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# Report of the Working Group on Widely Distributed Stocks (WGWIDE) 

28 August- 3 September 2018<br>Torshavn, Faroe Islands

# International Council for the Exploration of the Sea Conseil International pour l'Exploration de la Mer 

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## Executive Summary

The Working Group on Widely Distributed Stocks (WGWIDE) met in Tórshavn, Faroe Islands, during 28 August3 September 2018. The meeting, chaired by Guðmundur J. Óskarsson, was attended by 31 delegates and 5 by correspondence from 14 countries. The WG reports on the status and considerations for management of Northeast-Atlantic mackerel, blue whiting, Western and North Sea horse mackerel, Northeast-Atlantic boarfish, Norwegian spring-spawning herring, striped red mullet (Subareas 6, 8 and Divisions 7.a-c, e-k and 9.a), and red gurnard (Subareas 3, 4, 5, 6, 7, and 8) stocks. Additionally, a special request from the European Commission on interarea flexibility of horse mackerel fishery was addressed.

Northeast-Atlantic (NEA) Mackerel. This species is widely distributed throughout the ICES area and currently supports one of the most valuable European fisheries. Mackerel is fished by a variety of fleets from many countries (ranging from open boats using handlines on the Iberian coasts to large freezer trawlers and Refrigerated Sea Water (RSW) vessels in the Northern Area). The stock was benchmark in 2017 and the 2018 assessment was an update assessment, incorporating a new year for the catch information, for the IESSNS survey and for the RFID tagging recapture data (no new egg survey and recruitment index not available). The 2018 assessment revises the stock downward, and indicates that the SSB has been declining continuously since 2011, while the fishing mortality has been increasing. SSB in 2018 is estimated to be below MSY $B_{\text {trigger }}$ and $F$ larger that $\mathrm{F}_{\mathrm{pa}}$, which represents a deterioration of stock status compared to last year.

Blue Whiting. This pelagic gadoid is widely distributed in the eastern part of the North Atlantic. The assessment this year followed the Stock Annex based the conclusions from the Inter-Benchmark Protocol of Blue Whiting (IBPBLW 2016). Most of the annual catches are taken in the first half-year, which makes it possible to use preliminary catches for 2018 in the assessment. This is done to reduce the effect of potential biases from the single survey used for this assessment. The SSB of the stock is large but declining since 2017. F has been reduced in recent years, but is still above Fmsy. Recruitments in 2017 and 2018 are estimated to be low, following a period of high recruitments.

Western Horse Mackerel. This species is widely distributed throughout the Northeast Atlantic: it spawns in the Bay of Biscay, and in UK and Irish waters; after spawning, parts of the stock migrate northwards into the Norwegian Sea and the North Sea. The stock is assessed using the Stock Synthesis integrated assessment model. The 2018 assessment is an update of the benchmark assessment with the inclusion of the 2017 data. According to the assessment results, the 20152017 recruitment estimates are the highest observed since 2008 (and higher than the geometric mean estimated over the years 19832017). Fishing mortality since 2012 has been decreasing, dropping to low values in 20152017 due to lower catches and a reduced proportion of fraction of the adult population in the exploited stock; it is currently below FMSY. SSB in 2017 is estimated as the lowest in the time-series, below the precautionary reference point but above the limit reference point. The updated assessment shows the same trend as the previous ones, but rescales the absolute level of SSB and F over the most recent decade and, although this years' revision is smaller, this indicates that there is still considerable uncertainty associated with it. An inter-benchmark workshop has been scheduled for 2019: the workshop will aim at the revision of the biomass reference points and at investigate the causes of the instability in the assessment.

North Sea Horse Mackerel. After being benchmarked in January 2017, the CGFS and NS-IBTS survey indices were modelled with a zero inflated model to produce a combined index. The observed trend in the last years suggest that the stock is still at low levels in comparison with values in the early time-series. In 2017, the survey index shows a steep decline in comparison with year 2016. Despite this abrupt change in the survey abundance index, the catch advice for 2019 (decided in 2017) was not modified. The result of Length Based Methods to estimate proxy MSY reference points for the North Sea Horse Mackerel indicate that in 2016 and 2017 fishing mortality was slightly above Fmsy.

Northeast Atlantic Boarfish. This is a small, pelagic, planktivorous, shoaling species, found at depths of 0 to 600 m . The species is widely distributed from Norway to Senegal. The directed fishery for boarfish in the NEA is a relatively new one with large catches during the early 2000s when the fishery was unregulated. Catches have reduced significantly since 2012 to the current level. Annual catch advice is provided using the data limited category 3 approach based on output from an exploratory Bayesian surplus production assessment model. The assessment model utilises catch data, an acoustic survey estimate of stock size and indices from a number of bottom-trawl surveys. The current assessment indicates that since a historic high in 2012 biomass has declined sharply to a stable and low level since 2014.

Norwegian Spring Spawning Herring. This is one of the largest herring stocks in the world. It is highly migratory and distributed throughout large parts of the NE Atlantic. This stock was benchmarked in 2016 (WKPELA). The assessment model introduced in the benchmark (XSAM), incorporates uncertainty in the input data, and has been used to provide advice after the benchmark. The SSB on 1 January 2018 is estimated by XSAM to be above $\mathrm{B}_{\mathrm{pa}}$ ( 3.184 million t ). The stock is declining and the SSB time-series from the 2018 assessment is in line with the SSB time-series from the 2017 assessment. Fishing mortality in 2017 is estimated to be above the management plan $F$ that was used to give advice for 2017. A new management plan is being developed for the 2019 advisory year

Striped Red Mullet in North Sea, Bay of Biscay, Southern Celtic Seas, Atlantic Iberian Waters. 2016 was the first year this stock was considered by WGWIDE. This is a category 5 stock without information on abundance or exploitation, and the evaluation is based on commercial landings. The advice for this stock was given last year for 2018, 2019 and 2020.

Northeast-Atlantic Red Gurnard. 2016 was the first year this stock was been considered by WGWIDE. This is a category 6 stock for which there is no indication of where fishing mortality is relative to proxies and no stock indicators, and the evaluation is based on commercial landings. The advice for this stock was given last year for 2018 and 2019.

### 1.1 Terms of References (ToRs)

The Working Group on Widely Distributed Stocks (WGWIDE), chaired by Guðmundur J. Óskarsson, Iceland, met at The Faroe Marine Research Institute, Tórshavn, Faroe Islands 28 August - 3 September 2018 to:
a) Address generic ToRs for Regional and Species Working Groups;
b) Estimate MSY proxy reference points for the category 3 stocks in need of new advice in 2018.
c) Address a special request from the European Commission on interarea flexibility in catches for horse mackerel in divisions 8c and 9a.

### 1.1.1 The WG work 2018 in relation to the ToRs

With respect to ToR a, WGWIDE adopted the assessments of all the eight stocks, which formed the basis for stock status and the premise for the forecasts and advice. Based on the assessments the group produced a draft advice on TAC for four of the stocks, while a multi-annual advice from 2017 was in force for the other four (boarfish, red gurnard North Sea horse mackerel and striped red mullet). The individual stock report sections were not reviewed in plenary due to time constraints but audited by WG members right after the meeting. The summary sheets for all stocks were reviewed and agreed upon in plenary.
ToR b did not apply to any stock within the WG this year. This was because ICES gave a multi-annual advice on fishing opportunities for the two stocks in category 3 (boarfish and North Sea horse mackerel) in 2017. Next advice on these stocks will be given in 2019.

The progress on the work to address ToR c was introduced to the WG and discussed. Considering the time constrain to answer the request, a multispecies model was not considered feasible approach. In the absence of recent MSE work for the western stock, SimpSIM was selected for testing the impact of the flexibility for that stock while the southern stock has already a fully developed MSE approach. Some preliminary results were presented, but some work is still needed, including compiling the results with the southern stock results. It will be done in the coming weeks. In the WG it was pointed out that since $\mathrm{Fmsy}^{\text {is }}$ limited by $\mathrm{F}_{\mathrm{pa}}$ for the western stock, fishing above an advice based on FMSY will not be in accordance with a precautionary approach, which gives a short answer to the request.

### 1.2 Participants at the meeting

WGWIDE 2018 was attended by 31 delegates from the Netherlands, Ireland, Spain, Norway, Portugal, Iceland, UK (England and Scotland), Faroe Islands, Denmark, Greenland, and Russia. Five other fisheries scientists participated by correspondence. The full list of participants is in Annex 1.

### 1.3 Overview of stocks within the WG

Currently there are eight widely distributed and highly migratory stocks assessed in the WG with different methods, as indicted in the table below:

| Stock | ICES CODE | CATEGORY | ASSESSMENT METHOD |
| :--- | :--- | :---: | :--- |
| Boarfish | Boc.27.6-8 | 3 | Fproxy multiplier/ DLS category 3 |
| Red gurnard | Gur.27.3-8 | 6 | Qualitative evaluation |
| Norwegian <br> spring-sp. <br> herring | Her.27.1-24a514a | 1 | XSAM |
| Western horse <br> mackerel | Hom.27.2a4a5b6a7a- <br> ce-k8 | 1 | Stock Synthesis (SS) |
| North Sea horse <br> mackerel | Hom.27.3a4bc7d | 3 | Fproxy multiplier/ DLS category 3.1.0 |
| NE-Atlantic <br> mackerel | Mac.27.nea | 1 | SAM |
| Striped red <br> mullet | Mur.27.67a-ce-k89a | 5 | Qualitative evaluation |
| Blue whiting | Whb.27.1-91214 | 1 | SAM |

### 1.4 Quality and Adequacy of fishery and sampling data

### 1.4.1 Sampling Data from Commercial Fishery

The working group again carried out a brief review of the sampling data and the level of sampling on the commercial fisheries. Details are given in the relevant sections of this report.

Information on sampling data on the species newly included into WGWIDE, boarfish (Caprosaper), Striped red mullet (Mullussurmuletus) and Red Gurnard (Chelidonichthyscuculus) are also given in the relevant sections.

Length frequency data on gurnards are available from France and Spain, meaning that approximately two thirds of total landings are sampled. Provision of length data by métier from other countries would improve the understanding of exploitation patterns for this species.

Given the high value of striped red mullet, sampling and aging opportunities have been limited. The patchy distribution of the species and the noisy survey data limits the usefulness of these fishery-independent data sources. Further efforts should be made to improve precision of cpue series from French reference fleets as an indicator of stock status.

In general, to facilitate age-structured assessment, samples should be obtained from all countries with catches of the relevant species.

### 1.4.2 Catch Data

The WG has on a number of occasions discussed the accuracy of the catch statistics and the possibility of large scale under reporting or species and area misreporting.
The working group considers that the best estimates of catch it can produce are likely to be underestimates.

A specific issue on the catch data was reported to the WG. An issue on species allocation of catches exists for red gurnard. Before 1977, red gurnard was not specifically reported. Still, gurnards are not always reported by species, but rather as mixed gurnards. This makes interpretations of the records of official landings difficult and needs to be improved.

### 1.4.3 Discards

From 2015 onwards a landing obligation for European Union fisheries was introduced for fisheries directed on small pelagic fish including mackerel, horse mackerel, blue whiting and herring. However, as the landing obligation is introduced stepwise by fisheries at present discarding of small pelagic species can still legally occur in other fisheries. A general discard ban is already in place for Norwegian, Faroese and Icelandic fisheries.

Historically discarding in pelagic fisheries was more sporadic than in demersal fisheries. This is because the nature of pelagic fishing is to pursue schooling fish, creating hauls with low diversity of species and sizes. Consequently, discard rates typically show extreme fluctuation ( $100 \%$ or zero discards). High discard rates occurred especially during 'slippage' events, when the entire catch is released. The main reasons for 'slipping' are daily or total quota limitations, illegal size and mixture with unmarketable bycatch. Quantifying such discards at a population level is extremely difficult as they vary considerably between years, seasons, species targeted and geographical region.
Discard estimates of pelagic species from pelagic and demersal fisheries have been published by several authors. Discard percentages of pelagic species from demersal fisheries were estimated between $3 \%$ to $7 \%$ (Borges et al., 2005) of the total catch in weight, while from pelagic fisheries were estimated between $1 \%$ to $17 \%$ (Pierce et al. 2002; Hofstede and Dickey-Collas 2006, Dickey-Collas \& van Helmond 2007, Ulleweit \&Panten 2007, Borges et al. 2008, van Helmond et al. 2009, 2010, 2011, van Overzeeet al. 2013, Ulleweit et al. 2016). Slipping estimates have been published for the Dutch freezer trawler fleet only, with values at around $10 \%$ by number (Borges et al. 2008) and around $2 \%$ in weight (van Helmond et al. 2009, 2010 and 2011) over the period 2003-2010. Nevertheless, the majority of these estimates were associated with very large variances and composition estimates of 'slippages' are liable to strong biases and are therefore open to criticism.
Because of the potential importance of significant discarding levels on pelagic species assessments, the Working Group again recommends that observers should be placed on board vessels in those areas in which discarding occurs, and existing observer programmes should be continued. Furthermore, agreement should be made on sampling methods and raising procedures to allow comparisons and merging of dataset for assessment purposes. The newest update on discards for the different stocks assessed by the WG is provided in the sections for each of the stocks.

### 1.4.4 Age-reading

Reliable age data are an important prerequisite in the stock assessment process. The accuracy and precision of these data, for the various species, is kept under constant review by the Working Group. The newest updates on this aspect for the different stocks are addressed below.

### 1.4.4.1 Mackerel

The last otolith exchange was carried out in 2013/2014 by TI-SF. In order to increase the agreement between the laboratories involved in stock assessment especially for older fish, a workshop on age estimation of Atlantic Mackerel (Scomberscombrus) is scheduled for October 2018.

The sensitivity of the mackerel assessment to the effect of ageing errors on the input data to the assessment was investigated in the Workshop on Mackerel biological Quality Indicators (WKMACQI, ICES 2018c). An ageing error matrix was first derived from an otolith exchange workshop conducted in 2010 (ICES, 2010). The approach taken by the group was to start from the assumption that data currently used for the assessment were not affected by ageing errors. The error matrix was used to "pollute" the input data structured by age (catch-at-age, survey-at-age, weights-at-age, proportion ma-ture-at-age) and the assessment run on these data. Results show that the estimated stock trajectory in the recent year is very sensitive to the effect of ageing errors on input data ( $+12 \%$ for the SSB and $-17 \%$ for $\mathrm{F}_{\mathrm{bar}}$ ), specifically those used in model fitting (catches and survey at age). Changes in these data result in different estimated parameters (leading to a slightly different weighting of the difference data sources). Ageing errors therefore appear to be an additional source of uncertainty in the mackerel assessment that has not been considered so far.

### 1.4.4.2 Horse mackerel

Following the workshop in 2012 and the exchange in 2015 the last workshop on age reading of Trachurus trachurus, T. mediterraneus and T. picturatus was carried out in October 2015.

The workshop achieved quite a lot in terms of overcoming some of the major difficulties in ageing otoliths of Trachurus species. The results of the comparison between different ageing techniques on the same set of fish, showed a bias between readers and so it is recommended to use only one ageing technique by each reader. Moreover, the precision of reading is the same between slices and whole otoliths and so there is not one best ageing technique for T. trachurus. The progress of reading showed a percentage of agreement close to $65 \%$ for T. trachurus.

The next workshop on age reading of horse mackerel, Mediterranean horse mackerel and blue jack mackerel is scheduled for the beginning of November 2018.

### 1.4.4.3 Norwegian Spring-spawning Herring

A workshop on age reading of Norwegian Spring-spawning herring was carried out in November 2015. The meeting was attended by 12 experts from four countries. The workshop was a request from WGWIDE to WGBIOP to review any technical problems regarding age-reading of Norwegian spring-spawning herring between Norway, Denmark, Iceland and the Faroe Islands. This workshop was initiated after the IESNS survey in 2015, because there were concerns regarding dissimilarities between the age distributions from the different nations.

The workshop concluded that the different ages obtained from scale and otolith readings could be due to a number of issues relating to identification of the first winter ring and age interpretation of older fish, additionally confounded by stock mixing issues. Final conclusions could not be reached based on the samples from this workshop. With regards to the issue with sampling methods, WKPELA in March 2016 concluded that in general the biological samples are representative with regards to length distribution of NSS herring in the IESNS survey.
Therefore, it was recommended, that a follow up Workshop on Age estimation on Norwegian spring-spawning herring should consider the short-comings of the 2015 workshop and develop an ageing protocol that contains robust procedures for a quality check. The ageing issues should be addressed in full based on a larger sample set of
good quality scales and otoliths from the same fish and defined instructions for annotation. Prior to the follow up workshop within-country disagreements need to be resolved. Also, stock mixing issues need to be addressed (potentially by genetics combined with otolith shape analysis) and sampling protocols need to ensure that both otoliths and scales are sampled from the same fish (at least subsamples). This workshop has not yet been held, but a scale- and otolith exchange has taken place, and WGWIDE group members recommend that the workshop is held in the winter 2018/19.

### 1.4.4.4 Blue Whiting

The last workshop on age reading of blue whiting (WKARBLUE2) took place this year (2017), in June. This workshop was preceded by an otolith exchange, which was undertaken using WebGR in the year prior to the workshop. The actual otoliths were also sent around to all participating institutes. The exchanged otolith collection included 245 images. The overall agreement with modal age of the pre-workshop exercise was $64.1 \%$ considering all readers and $70 \%$ for the assessment readers. During the workshop 129 otoliths with annotations were discussed in plenary and $85 \%$ agreement was achieved. There were no clear signs of seasonal misinterpretations, but the Mediterranean and most northern areas (ICES area 27.14.b and NAFO 1C) proved to be quite difficult.
Different methods to help age readers determining a zone was discussed during the workshop. The burning of otoliths showed some potential in interpreting the inner ring, but not to be used as a routine. The sliced technique besides being time consuming do not show advantages on ring interpretation, and in turn can also introduces more misinterpretation on ageing. During the workshop some of the otoliths from the exercise were polished, to help readers in the cases were the age rings were not so evident, completely absent, or showing a growth pattern different from the expected. The polishment results revealed to be useful on the ring interpretation and to help during the plenary discussion, although we do not recommend this technique to be used as a routinely procedure, as it is very time consuming. Plug-in for ImageJ (OtoRing), which can detect variation in opacity in the otolith surface and be used as a tool on age rings identification as presented (Gonçalves et al. 2017a). Furthermore, a criteria table with possible otolith ring diameters from an IPMA study was tested during the workshop. The table showed potential, but a larger dataset is still needed before it can be implemented as a guideline. This dataset will consider samples by area and sex to achieve criteria's classification which take into account those differences in growth patterns, due to the blue whiting sexual dimorphism (Gonçalves et al. 2017b).
A study on the otoliths from the Portuguese coast showed differences between the first ring length in this area and the described in the literature ( 8.33 and 9.33 mm ). Rings measurements of the first annulus, taken during the workshop, revealed also differences between ICES areas (27.2.a - 27.9.a), 27.14.b and Mediterranean.
The reoccurring problems among age readers were identification of the position of the first annual growth ring, false rings and interpretation of the edge. In order to outcome those problems, age validations studies on blue whiting otoliths were further recommended and should be conducted until the next age reading workshop. The next workshop on the age estimation of blue whiting will be carried out in June 2020. An exchange on age reading calibration was in preparation and is planning to start at the beginning of 2019.

### 1.4.4.5 Boarfish

This stock will be part of the EU data collection framework from 2017 onwards. Age length keys were produced in 2012. The age reading was conducted by DTU Aqua on samples from all three countries in the fishery: Ireland, Denmark and UK (Scotland).

### 1.4.4.6 Striped red mullet

In 2011, an Otolith Exchange Scheme has been realized, which was the second exercise for the striped red mullet. For details see section 12.7.

### 1.4.4.7 Red gurnard

Age data were available for red gurnards from the EVHOE and IGFS surveys. Understanding of this stock would be improved by reading otoliths from other surveys in the assessment area (e.g. NS-IBTS, SCO-WCS, CGFS) which contribute to perceptions of red gurnard stock status in terms of their cpue series.

### 1.5 Quality Control and Data Archiving

### 1.5.1 Current methods of compiling fisheries assessment data

Information on official, area misreported, unallocated, discarded and sampled catches have again this year been recorded by the national laboratories on the WG-data exchange sheet (MS Excel; for definitions see text table below) and sent to the stock coordinators and uploaded through the InterCatch hosted application. Co-ordinators collate data using the either the sallocl (Patterson, 1998) application which produces a standard output file (Sam.out) or the InterCatch hosted application.

There are at present no specified criteria on the selection of samples for allocation to unsampled catches. The following general process is implemented by the species coordinators. A search is made for appropriate samples by gear (fleet), area, and quarter. If an exact match is not available the search will extend to adjacent areas, should the fishery extend to this area in the same quarter. Should multiple samples be available, more than one sample may be allocated to the unsampled catch. A straight mean or weighted mean (by number of samples, aged or measured fish) of the observations may be used. If there are no samples available the search will move to the closest nonadjacent area by gear (fleet) and quarter, but not in all cases.

It is not possible to formulate a generic method for the allocation of samples to unsampled catches for all stocks considered by WGWIDE. However full documentation of any allocations made are stored each year in the data archives (see below). It should be noted that when samples are allocated the quality of the samples may not be examined (i.e. numbers aged) and that allocations may be made notwithstanding this. The Working Group again encourages national data submitters to provide an indication of what data could be used as representative of their unsampled catches.

Following the introduction of the landings obligations for EU fisheries new catch categories had to be introduced from 2015 onwards. The catch categories used by the WGWIDE are detailed below:

| Official Catch | CATCHES AS REPORTED by the official statistics to ICES |
| :--- | :--- |
| Unallocated Catch | Adjustments (positive or negative) to the official catches made for any <br> special knowledge about the fishery, such as under- or over-reporting <br> for which there is firm external evidence. |
| Area misreported Catch | To be used only to adjust official catches which have been reported <br> from the wrong area (can be negative). For any country the sum of all <br> the area misreported catches should be zero. |
| BMS landing | Landings of fish below minimum landing size according to landing <br> obligation |
| Logbool registered <br> discards | Discards which are registered in the logbooks according to landing <br> obligation |
| Discarded Catch | Catch which is discarded |
| WG Catch | The sum of the 6 categories above |
| Sampled Catch | The catch corresponding to the age distribution |

### 1.5.2 Quality of the Input data

Primary responsibility for the accuracy of national biological data lies with the national laboratories that submit such data. Each stock co-ordinator is responsible for combining, collating, and interpolating the national data where necessary to produce the input data for the assessments. A number of validation checks are already incorporated in the data submission spreadsheet currently in use, and these are checked by the coordinators who in the first instance report anomalies to the laboratory which provided the data.

The working group acknowledges the effort some members have made to provide "corrected" data, which in some cases differ significantly from the officially reported catches. Most of this valuable information is gathered on the basis of personal knowledge of the fishery and good relations between the responsible scientist and the fishermen. The WG is aware of the problem that this knowledge may be lost if the scientist resigns, and asks the national laboratories to ensure continuity in data provision. In addition, the working group recognises and would like to highlight the inherent conflict of interest in obtaining details of unallocated catches by country and increasing the transparency of data handling by the Working Group.

Overall, data quality has improved and sampling deficiencies have been reduced compared to earlier years, partly due to the implementation of the EU sampling regulation for commercial catch data. However, some nations have still not or inadequately aged samples. Occasionally, no data are submitted such that only catch data from EuroStat is available, which are not aggregated quarterly but are yearly catch data per area.

The Working Group documents sampling coverage of the catches in two ways. National sampling effort is tabulated against official catches of the corresponding country (see stock specific sections). Furthermore, tables showing total catch in relation to numbers of aged and measured fish by area give a picture of the quality of the overall sampling programme in relation to where the fisheries are taking place. These tables are contained in the species sections of this report.

The national data on the amount and the structure of catches and effort are archived in the ICES Intercatch database. The data are provided directly by the individual countries and are highly aggregated for the use of stock assessments.

There exist gaps in some dataseries, in particular for historical periods. The WG has requested members to provide any national data reported to previous working groups
(official catches, working group catches, catch-at-age and biological sampling data) not currently available to the WG. Furthermore, the WG recommends that national institutes increase national efforts to collate historic data.

### 1.5.3 Stock data problems relevant to data collections

A number of other stock data problems were brought forward to the contact person and are listed in table below for the information of ICES-Working Groups and RCMs as specified.

| Stock | Data Problem | How to be AdDressed in | BY wнo |
| :---: | :---: | :---: | :---: |
| Northeast <br> Atlantic <br> Mackerel | Submission of data | Data submissions must include all the data outlined in the data call and be submitted by the deadline. <br> Should the data submitter be unavailable after the data has been submitted (e.g. vacation) an alternative contact should be available who can be contacted in the event of any queries. | National laboratories |
| Northeast <br> Atlantic <br> Mackerel | Discard and slippage information | Discard and slippage information is incomplete. All fleets should be monitored and sampled for discard and slipping. Data should be supplied to the coordinator by the submission deadline, accompanied by documentation describing the sampling protocol. | National laboratories, RCG NA, RCG NS\&EA |
| Northeast <br> Atlantic <br> Mackerel | Sampling deficiencies- general | All countries involved should provide sampling information. Increased cooperation between countries would help reduce redundancy and increase coverage. | National laboratories, RCG NA, RCG NS\&EA |
| Northeast <br> Atlantic <br> Mackerel | Sampling of foreign vessels | Any information available from the sampling of foreign vessels should be forwarded to the appropriate person in the national laboratory in order that they may use this information when compiling the data submission. | National laboratories; RCG NA, RCG <br> NS\&EA |
| Boarfish | Boarfish only measured to the 1 cm on the IBTS by some countries | Following the MoU between ICES and EU, boarfish (Capros aper) was included into WGWIDE. Boarfish should be measured to the 0.5 cm on the IBTS due to the small length range and the relatively high ages observed. | ICES IBTSWG |


| Stock | Data Problem | How to be Addressed in | BY who |
| :---: | :---: | :---: | :---: |
| Horse <br> Mackerel - <br> Western <br> Stock | Uncertainties in the use of the current egg production method for the assessment | Investigate spawning biology and investigate potential methods to incorporate time varying fecundity in the assessment. | Future Benchmark |
| Horse <br> Mackerel - <br> Western <br> Stock | Assumed value of 0.15 for M. | Value of 0.15 should be investigated. | Future Benchmark |
| Horse <br> Mackerel - <br> Western <br> Stock | No ageing error included in the assessment. | Different values for ageing error should be investigated. | Future Benchmark |
| Horse <br> Mackerel - <br> Western <br> Stock | Partial information from acoustic survey in division 8 . | Information from all the surveys carried out in the area should be pulled together. | Future Benchmark |
| Horse <br> Mackerel - <br> North Sea <br> Stock | Incomplete report of discards by nonpelagic fleet. | Reporting of discards by national institutes. | National Institutes |
| Horse <br> Mackerel - <br> North Sea <br> Stock | Low level of sampling and survey data. Currently IBTS and CGFS data are available. | Collection of information from other working groups. Possible implementation of an acoustic survey for horse mackerel in 3rd or 4th Quarter. | WGBIOP, WGCATCH, RCG NS\&EA |
| Horse <br> Mackerel - <br> North Sea <br> Stock | Lack of maturity ogive both by age or length | Collection of information about maturity stage during regular biological sampling (otoliths) in commercial and survey fleets | National institutes |
| Horse <br> Mackerel - <br> North Sea <br> Stock | Absence of length distribution in the discard component | Sampling of length distribution of discarded individuals | National institutes |
| Horse <br> Mackerel - <br> North Sea <br> Stock | Low contribution of countries to the estimation of the age and length distribution of catches | Sampling of age and length information from commercial with a distribution of sampling effort over the year and areas in the North Sea | National institutes |
| Norwegian <br> Spring- <br> spawning <br> Herring | Low sampling effort on some nations (considerably lower than the 1 sample/1000 tonnes recommended for this stock by EU) | Sampling effort should be increased by nations with little or no samples. | National laboratories; RCG NS\&EA |


| Stock | Data Problem | How to be Addressed in | Br who |
| :---: | :---: | :---: | :---: |
| Northeast <br> Atlantic <br> Blue <br> Whiting | Submission of data | Data submissions must inlcude all the data outlined in the data call and be submitted by the deadline. Should the data submitter be unavailable after the data has been submitted (e.g. vacation) an alternative contact should be available who can be contacted in the event of any queries. | National laboratories |
| Red gurnard | Discard and slippage information | Discard rates for this species can be very high (up to $100 \%$ of catch at a trip level). Alternative data sources and methods for estimation (e.g. CCTV systems) should be investigated. | National laboratories |
| Red gurnard | Stock area | Red gurnard is found all along the Iberian continental shelf. There are no records of catches of red gurnards in SA5, and this area could be removed from the data call. |  |

### 1.6 Comment on update and benchmark assessments

Update assessments were presented to the WG for all the eight stocks in the group. Western and North Sea horse mackerel, and NEA Mackerel were assessed on basis of benchmark that took place in January 2017 (ICES 2017a). In same way, blue whiting and Norwegian spring-spawning herring were assessed by the latest benchmarks (ICES 2016b and ICES 2016f, respectively). The other three stocks addressed by the WG have not been benchmarked recently but were still assessed by the WG. Result from the assessment model for boarfish is used as indicator of trends in the stock development. The catch data were updated for Red gurnard (Chelidonichthyscuculus) in Subareas 3-8 (Northeast Atlantic) and Striped red mullet (Mullussurmuletus) in Subareas 6 and 8 and Divisions 7.a-c, e-k and 9.a (West of Scotland, Bay of Biscay, Southern Celtic Seas, Atlantic Iberian Waters).

### 1.7 Latest benchmark results

None of the eight fish stocks within the WG have been taken to benchmark assessment since presented in the last year's report.

### 1.8 Planning future benchmarks

While five of the major stocks within the group has been benchmarked recently (20162017), boarfish has not been benchmarked yet at all, and there is a need for a benchmarked assessment. The WG propose that a benchmark for boarfish could take place in 2020 and have made an issue list to be addressed to support it. The WG discussed also if red gurnard was potential candidate for benchmark assessment in 2020. The conclusion was that it was not realistic, especially considering lack of essential information on stock's identity, mixing and distribution.

During the WGWIDE meeting, several potentially serious problems were identified in the analytical assessment of mackerel. Consequently, a list of issues specifically related to this was made to support an inter-benchmark to take place as soon as possible.
In the same way, an issue list was made to support inter-benchmark for western horse mackerel to take place as soon as possible. The issues are related to the reference points for the stock and to a lack of an alternative assessment model for a comparison.

### 1.9 Special Requests to ICES regarding stocks within WGWIDE

Two requests were directed to WGWIDE, on evaluation of a long-term management strategy for Norwegian spring-spawning herring (addressed by WKNSSHMSE; ICES 2018b), and on Horse mackerel in areas 8c and 9a - interarea flexibility (addressed by WGWIDE, this report).

### 1.9.1 Request to ICES from NEAFC concerning a long-term management strategy for Norwegian spring-spawning herring.

In order to revise the long-term management plan for Norwegian spring-spawning herring consistent with the new stock assessment model (ICES 2016; 2017) and the corresponding updated reference points (ICES 2018a; 2018b), a Management Strategy Evaluation is needed. The objective is to ensure harvest of the stock within safe biological limits. The Parties therefore request ICES to evaluate the following harvest control rules.

## Rule 1

- A range of $B_{\text {trigger }}$ from 1 to 6 million tonnes with a range of target Fs from 0.05 to 0.25 .
- The fishing mortality is the average for age groups 5 to $12+$ weighted by stock numbers.
- Time of comparison for SSB is the same as used in the assessment.
- A harvest control rule with a fishing mortality equal to the target F when SSB is at or above $B_{\text {trigger }}$.
- In the case that the SSB is forecast to be less than $B_{\text {trigger, }}$ the TAC shall be fixed consistently with a fishing mortality that is given by: $F=F_{\text {target }} * S S B / B_{\text {trigger }}$
- The following special case is to be evaluated: $B_{\text {trigger }}=3.184\left(=M S Y B_{\text {trigger }}=B_{p a}\right)$ and the target fishing mortality of $0.102\left(F_{M S Y}\right)$.
Rule 2
- A range of $B_{\text {trigger }}$ from 2.5 to 6 million tonnes with a range of target Fs from 0.05 to 0.25 .
- The fishing mortality is the average for age groups 5 to 12+ weighted by stock numbers.
- Time of comparison for SSB is the same as used in the assessment.
- A harvest control rule with a fishing mortality equal to the target $F$ when SSB is at or above $B_{\text {trigger }}$.
- In the case that the SSB is forecast to be less than $B_{l i m}$, the target $F$ is 0.05.
- In the case that the $S S B$ is forecast to be between $B_{l i m}$ and $B_{\text {trigger, }}$ the target $F$ will decrease linearly between those two points.
- The following special case is to be evaluated: $B_{\text {trigger }}=3.184\left(=M S Y B_{\text {trigger }}=B_{p a}\right)$ and the target fishing mortality of $0.102\left(F_{M S Y}\right)$.

Rule 3

- A proxy for $\operatorname{SSB}\left(S S B_{p r o x y}\right)$ is defined as the biomass of herring aged 5 and older or an appropriate age range as identified by ICES.
- The reference biomass $\left(B_{r e f}\right)$ is defined as the biomass of herring aged 4 and older or an appropriate age range as identified by ICES.
- Time of comparison for $S S B_{\text {proxy }}$ is the same as used for SSB in the assessment.
- A range of $B_{\text {trigger }}$ from 1 to 6 million tonnes with an appropriate range of harvest rate ( $H_{R_{\text {target }}}$ ).
- A harvest control rule with TAC $=H R_{\text {target }} * B_{\text {ref }}$ when $S S B_{\text {proxy }}$ is at or above $B_{\text {trigger }}$.
- In the case that the $S S B_{\text {proxy }}$ is forecast to be less than $B_{\text {trigger, }}$ the $T A C=H R_{\text {target }} * B_{\text {ref }} *$ $\left(S S B_{\text {proxy }} / B_{\text {trigger }}\right)$
- The following special case is to be evaluated: Btrigger $=3.184\left(=M S Y B_{\text {trigger }}=B_{p a}\right)$ and a harvest rate equivalent to $0.102\left(F_{M S Y}\right)$.

Rule 4
A biomass rule intended to be equivalent to Rule 2 with two levels of harvest rate: target harvest rate $=H R_{\text {target }}$ when SSBproxy is greater than Btrigger; harvest rate $=H R_{\text {lowest }}$ when SSBproxy $^{\text {is below }}$ $B_{\text {lim; }}$ and harvest rate decreasing linearly between these bounds.

## Evaluation and performance criteria

Starting point of the evaluations should be the current stock status as estimated by the most recent assessment and be consistent across time.
Each alternative shall be assessed in relation to how it performs in the short term (2019-2023), medium term (2024-2033) and long term (2034-2053) in relation to:

- Average SSB
- Average yield
- Indicator for year-to-year variability of SSB and yield
- Risk of SSB falling below Blim


## Evaluation of the management strategies shall be simulated:

- With no constraint on the interannual variation of TAC.
- With a constraint on the interannual variation of TAC:
- When the rules would lead to a TAC, which deviates by more than $20 \%$ below or $25 \%$ above the TAC of the preceding year, the TAC is to be set respectively no more than $20 \%$ less or $25 \%$ more than the TAC of the preceding year.
- The TAC is to be set as the average of a) the current TAC and b) the TAC that would result from the application of the harvest control rule without constraint for the TAC year.
- The TAC constraint shall not apply if the SSB (rule 1 and 2) or SSBproxy (rule 3 and 4) in the year for which the TAC is to be set is less or equal to Btrigger.
- Allowing a maximum of $10 \%$ to be banked or borrowed any year.

ICES is also requested to assess what, if any, other measures in addition to those contained in the present Management Strategy might contribute to attaining the objectives of the strategy, and provide estimates of their efficiency.

Finally, it is expected that the Parties will, as appropriate, review and revise these management measures and strategies on the basis of any new advice provided by ICES.

## References:

ICES. 2016. Report of the Benchmark Workshop on Pelagic Stocks (WKPELA), 29 February- 4 March 2016, ICES Headquarters, Copenhagen, Denmark. ICES CM 2016/ACOM:34. 106 pp.

ICES. 2017. Report of the Working Group on Widely Distributed Stocks (WGWIDE), 30 August5 September 2017, ICES Headquarters, Copenhagen, Denmark. ICES CM 2017/ACOM:23. 994 pp.

ICES. 2018a. Workshop on the determination of reference points for Norwegian Spring-spawning (WKNSSHREF), 10-11 April 2018, ICES Headquarters, Copenhagen, Denmark. ICES CM 2018/ACOM:45. 83 pp.

ICES. 2018b. Special Request Advice Northeast Atlantic and Arctic Ocean Ecoregions, 26 April 2018 sr.2018.06
https://doi.org/10.17895/ices.pub. 4295
1.9.2 Special request to ICES from European Commission, DG MARE, C1 on an interarea flexibility between 8.C and 9.A for horse mackerel (Western and Southern stock).

Background: The Horse mackerel (Trachurustrachurus) stocks in Subarea 8 and divisions 2.a, 4.a, 5.b, 6.a, 7.a-c, and 7.e-k (the Northeast Atlantic) and in Division 9.a (Atlantic Iberian waters) are both classified by ICES as category 1 stock and apply the MSY approach. The TACs for horse mackerel will be set separately for 2018 as follows:


During the Council negotiations on TACs and quotas for 2018 the Commission received a request from Spain, asking for a change to the special condition by increasing the interarea flexibility from $5 \%$ to $15 \%$.

Request: To allow the Commission to consider a proposal for an amendment of the 2018 TAC regulation on the special condition applied to horse mackerel, ICES is requested to analyse any information deemed suited:
a ) to evaluate the impact of an increased interarea flexibility, from $5 \%$ to $15 \%$, to facilitate the implementation of the landing obligation, notably whether such an increase would be in line with the precautionary approach.
b) to evaluate what \% of interarea flexibility could be considered to be in line with the precautionary approach, if a negative opinion is given to a).

### 1.10 General stock trends for widely distributed and migratory pelagic fish species

This working group has carried out the stock assessments of the following widely distributed and migratory pelagic species: boarfish, red gurnard, Norwegian springspawning herring, Western horse mackerel, North Sea horse mackerel, Northeast Atlantic mackerel, Striped red mullet and Blue whiting.

Analytical (category 1) type of assessments are available for the four main species that make up the bulk of the biomass of pelagic species in the Northeast Atlantic:

- Northeast Atlantic mackerel
- Norwegian spring-spawning herring
- Blue whiting
- Western horse mackerel.

The fluctuations in the catches of the four stocks since 1988 are shown in Figure 1.10.1.

stockkeylabel ${ }^{\|}$her.27.1-24a514a \| hom.27.2a4a5b6a7a-ce-k8 \| mac.27.nea ${ }^{\|}$whb.27.1-91214

Figure 1.10.1: Catch of mackerel, western horse mackerel, blue whiting and Norwegian springspawning herring

The trends in SSB of the four stocks are shown in Figure 1.10.2, both in absolute biomass (tonnes) and in relative proportions. At the maximum, pelagic biomass of these species has been around 15 million tonnes. Recently the biomass appears to have decreased to around 12 million tonnes. The contributions of Norwegian Spring-spawning herring, Western horse mackerel and Northeast Atlantic mackerel has decreased in recent year while blue whiting has increased.

stockkeylabel ${ }^{\|}$her.27.1-24a514a ॥ hom.27.2a4a5b6a7a-ce-k8 ॥ mac.27.nea ${ }^{\|}$whb.27.1-91214

Figure 1.10.2: SSB of mackerel, western horse mackerel, blue whiting and Norwegian springspawning herring

An overview of the key variables for each of the stocks (Fishing pressure (F), recruitment (R) and Spawning-stock biomass (SSB)) is shown in Figure 1.10.3. From these comparisons it can be concluded that the fishing mortality of mackerel and blue whiting has generally been higher than the fishing mortality of horse mackerel and herring. Recruitment levels of blue whiting and herring are on a comparable scale and substantially higher and horse mackerel (except for the 1982 year class) and mackerel. Biomass
trends of the different stocks are somewhat on the same level but show very different tendencies.


Figure 1.10.3: SSB of mackerel, western horse mackerel, blue whiting and Norwegian springspawning herring

An overview of stock weight at age for mackerel and blue whiting is shown in figure 1.10.4. Older mackerel has experienced a substantially lower weight at in the recent years although this tendency appears to have changed for the younger ages. Weight at age of blue whiting shows substantial fluctuations which appear to be somewhat related to the stock size.


Figure 1.10.4 Trends in stock weight-at-age for mackerel and blue whiting.

### 1.11 Ecosystem considerations for widely distributed and migratory pelagic fish species

Number of studies demonstrate that environmental conditions (physical, chemical and biological) largely influence fish stocks productivity by changing the level of recruitment, growth rates, survival rates, or inducing variations in their geographical distribution (Skjoldal et al., 2004, Sherman and Skjoldal 2002). It has been acknowledged that future lines of work in stock assessment should take ecosystem considerations into account in order to reduce the levels of uncertainty regarding the present and future status of commercial stocks. Hence, WGWIDE encourages further work to be carried out on ecosystem considerations linked to widely distributed fish stocks including NEA mackerel, Norwegian spring-spawning herring, blue whiting and horse mackerel. A close collaboration with the Working Group on Integrated Assessment of Norwegian Sea (WGINOR; ICES 2018a), and hopefully others relevant Integrated Assessment groups within ICES in the near future, will help in operationalizing ecosystem approach for the widely distributed pelagics assessed in WGWIDE. The text below was largely provided by WGINOR (ICES 2016e; 2018).

### 1.11.1 Climate variability and climate change

The North Atlantic Oscillation (NAO) corresponds with the alternation of periods of strong and weak differences between Azores high and Icelandic low pressure centres. Variations in the NAO influence winter weather over the North Atlantic and has a strong impact on oceanic conditions (sea temperature and salinity, Gulf Stream intensity, wave height). The 2015 winter NAO index was high, and simultaneously cold/freshwaters on the Canadian site of the Atlantic that winter and spring because of increase advection resulted in relative low temperatures in the Sub Polar Gyre (SPG) and low temperatures at all depths in 2015 in the large part of the Northeast Atlantic in comparison to 20 years long-term mean (ICES, 2015). This positive NAO continued in 2016 and 2017.

The classical measure of global warming is the northern hemisphere Temperature anomaly (NHT) (Jones and Moberg, 2003) which is computed as the anomaly in the annual mean of seawater and land air surface temperature over the northern hemisphere. During the last three decades, NHT anomalies have exhibited a strong warming trend. Pelagic planktivorous species such as Northeast Atlantic mackerel (Astthorsson et al., 2012; ICES, 2013; Nøttestad et al. 2016), Norwegian spring-spawning herring and blue whiting may and have been taken advantage of warming oceans by extending their possible feeding opportunities further north, e.g. in Arctic waters. If such changes are, however, directly or indirectly driven by the warming are not fully understood (Olafsdóttir et al. 2018; Nikolioudakis et al.2018).

Acidification of the oceans is another event related to accumulation of anthropogenic greenhouse gases in the atmosphere. During the last 30 years, pH has decreased significantly in most water layers in Lofoten and the Norwegian basins. Different components like $\mathrm{CO}_{2}$, aragonite and number of other factors such as temperature, salinity, and alkalinity may affect pH and carbon systems in the ocean. The impacts of the acidification on the ecosystem remains to be explored.

### 1.11.2 Circulation pattern

The circulation of the North Atlantic Ocean is characterized by two large gyres: the Subpolar Gyre (SPG) and subtropical gyre (Rossby, 1999). When the SPG is strong it
extends far eastwards bringing cold and fresh Subarctic water masses to the NE Atlantic, while a stronger SPG allows warmer and more saline subtropical water to penetrate further northwards and westwards over the Rockall plateau area. Changes in the oceanic environment in the Porcupine/Rockall/Hatton areas have been shown to be linked to the strength of the Subpolar Gyre (Hátún et al., 2005). The large oceanographic anomalies in the Rockall region spread directly into the Nordic Seas, regulating the living conditions there as well as further south. Such changes are likely to have an impact on the spatial distribution of spawning and feeding grounds and on migration patterns of widely distributed pelagic fish species.

### 1.11.3 Recent trends in oceanography and zooplankton in Norwegian Sea

The time-series of ocean heat content in the Atlantic Water of the Norwegian Sea starting in 1951 show that the recent warm period continues (Figure 1.11.1). In fact, there is a continuing increase in 2017 compared to 2016 that was the previous record high value. This positive anomalous heat content is mainly confined to the Lofoten Basin. At the same time the freshwater content shows a slight increase, i.e. freshening, which is not expected from the usual T-S relation of warm/saline and cold/fresh varying in concert. Such change could be either due to anomalous air-sea exchange or changes in the volume or characteristics of the source water masses.

In the southern entrance to the Norwegian Sea there is a tendency toward slightly colder water compared to the recent years. However, more remarked is a prominent freshening trend pointing toward increased influence of water from the western Atlantic.

The index of Arctic water into the southern Norwegian Sea with the East Icelandic Current appears still weak compared to the condition previous to about 2002. This is further reflected in lower biomass of zooplankton in this region after around 2002 (Figure 1.11.2).

Upstream analysis of satellite sea surface height data indicates that the Subpolar gyre, which has been in a weak state for many years, has been strengthening during the last three years. If this trend continues, we should expect increased levels of silicate entering the Norwegian Sea over the coming years and consequently a reversal in the declining trend of silicate observed in the Norwegian Sea since 1990 (Rey 2012; Pacariz et al., 2016; Hátún et al., 2017). The atmospheric forcing represented by the winter NAOindex was positive, for the third consecutive winter. In the Labrador-Irminger Sea this normally means higher windstress curl and larger ocean to air heat loss and thus enhanced Subpolar gyre. For the Norwegian Sea, however, the averaged windstress curl showed low values, indicating that the atmospheric lows did not follow their normal route through the Norwegian Sea.


Figure 1.11.1. Time-series of anomalies of heat content of the Atlantic waters in Norwegian Sea (source:
http://www.imr.no/temasider/klima/klimastatus/norskehavet/norskehavet 2/nb-no).
The time-series of meso-zooplankton biomass in the Norwegian Sea from the International Ecosystem Survey in Norwegian Sea (IESNS) in May shows strong long-term variability (Figure 1.11.2). Following a maximum in biomass during the early 2000s the biomass declined with a minimum in 2006. From 2010 the downward trend reversed and reached back to the long-term mean again in 2014. Biomass dropped again in 2015 but have been increasing since then. Interestingly, all the areas, excluding east of Iceland and on few occasions Jan Mayen AF, show parallel changes in zooplankton biomass.


Figure 1.11.2. Indices of zooplankton dry weight ( $\mathrm{g} \mathrm{m}^{-2}$ ) sampled by WP2 in May in different areas in and near Norwegian Sea from 1995 to 2018 as derived from interpolation using objective analysis utilizing a Gaussian correlation function (ICES 2018b; see details on methods and areas in ICES 2016a).

### 1.11.4 Species interactions

The distribution of species considered by WGWIDE can overlap to a large extent during some part of the year, and density-dependent competition for food could be expected. All the species are potential predators on eggs and larvae and the larger species (mackerel and horse mackerel) are also potential predators of the juveniles. Consequently, cannibalism and interspecific predation is likely to play an important role in the dynamics of these pelagic stocks. As examples, density-dependent growth has been observed both for mackerel (Olafsdottir et al. 2015) and Norwegian spring-spawning herring (Hömrum et al. 2016). Furthermore, several studies on diet composition have shown a high overlap (see overview in ICES 2016a) and even intraguild predation between species, e.g. NEA mackerel predation on NSS herring larvae on the Norwegian shelf area (Skaret et al. 2015) and sardine predation on anchovy eggs in the Bay of Biscay (Bachiller et al. 2015).

The Norwegian Sea and adjacent waters are the main feeding ground for the three main small pelagic fish stocks (NSS herring and blue whiting in spring and summer) and NEA mackerel (in summer) (Skjoldal et al., 2004; Langøyet al. 2012; ICES 2018b). The three species are able to adapt their feeding strategy to different conditions, including preying in cold water masses, where they show significantly higher feeding incidence and stomach fullness (Bachiller et al. 2016). The increasing spatial overlap between herring and mackerel in recent years could be enhancing their interspecific competition, given the higher feeding efficiency of mackerel, as demonstrated by significantly higher stomach fullness indices, contrasting earlier periods with limited spatial overlap (Langøy et al. 2012, Debes et al. 2012, Oskarsson et al. 2015, Bachiller et al. 2016).

NEA mackerel and NSS herring share a similar diet based on calanoid copepods, especially C. finmarchicus, whereas blue whiting shows lower diet overlap with these two species, broader diet composition and dominance of larger prey like euphausiids and amphipods (Langøy et al. 2012, Bachiller et al. 2016). Recent estimates based on bioenergetics show that these three species consume on average 135 million tonnes of zooplankton per year (2005-2010; Bachiller et al. 2018), which are higher than previous estimates (e.g. Utne et al., 2012; Skjoldal et al., 2004). NEA mackerel consumed 23-38\%, NSS herring $38-51 \%$ and blue whiting $14-39 \%$ of the total zooplankton eaten by pelagic fish during the feeding season. This means that, in terms of consumption/biomass ratios, NEA mackerel feeding rates can be as high as that of the NSS herring during some years. Together, these three stocks were estimated to have consumed annually 53-81 million tonnes of copepods, 26-39 million tonnes of euphausiids and amphipods, 8-42 million tonnes appendicularians and 0.2-1 million tonnes of fish.

Sardine, mackerel, horse mackerel, blue whiting and herring have all been found in the diet of several cetacean and seabird species and are also part of the diet of other fish species (e.g. hake, tuna found with sardine and anchovy) (Anker-Nilssen and Lorentzen, 2004; Nøttestad et al. 2014). Comparison of population estimates of pelagic fish with those of top predators (e.g. minke whale, fin whale, killer whales) suggests that predation on pelagic fish by other pelagic fish has a much bigger potential for impact in regulating populations than that the predation by marine mammals and seabirds in the North Sea (Furness, 2002). Nevertheless, top predators could play a bigger role in pelagic fish dynamics at regional or local scales particularly when fish biomass is low (Nøttestad et al., 2004). In this WGWIDE report, several aspects of interaction between the pelagic fish stocks are addressed in the stock specific sections.

### 1.12 Future Research and Development Priorities

As part of the planning towards future benchmark assessments, the working group started in 2014 preparing a list of research priorities for each stock that can potentially improve the quality of the advice generated for each stock. This list is updated in every WG meeting, by removing issues as they have been solved and adding new ones when they arise. We have considered scientific research, improvements to data collection and development of assessment techniques, both generally and on a stock-by-stock basis, as appropriate. The most important of these developments are described below.

In general, more focus should be towards integrated ecosystem assessments for the stocks within WGWIDE. Some of the WGWIDE members also participate in the work of the Working Group on Integrated Assessment for Norwegian Sea (WGINOR), which help in communication between these two groups. However, there are also other regional Integrated Ecosystem Assessment groups that could be relevant to WGWIDE and the stocks assessed by it. We hope to put more emphasis on this in the coming years.

### 1.12.1 NEA Mackerel

Since the last benchmark of the mackerel assessment, the perception of the stock in the recent years has undergone several revisions. This was assumed to be a consequence of the short time-series (IESSNS and RDIF tagging data), whose catchability (survival rate for the tags) were not yet well estimated by the model. In addition, there was a conflicting signal between the egg survey and the IESSNS, and due to the changes in the relative weights of these surveys in the successive assessments (becoming more in favor of the egg survey), this would cause a downward revision of the stock size. At the WGWIDE 2018, it appeared that the sensitivity of the assessment to the tagging data had until now been overlooked, and that this data source (especially the RFID series) might be the most influential data source for this assessment. The tagging data are treated differently from conventional survey indices and the group lacks experience and proper diagnostics to understand its influence. Another problem with the RFID tags is that survival rate is estimated very low (ca. 0.1). This survival rate is a proxy for a number of factors: tagging mortality, scanning efficiency, tag loss and underestimation of the stock. The old steel tags indicated higher survival of 0.4. Furthermore, there is a strong pattern in the process error (model artificially adding fish to the stock) in the period where the stock and the catches increased rapidly, which also corresponds with the time the IESSNS and the RFID series started.

In order to get a better understanding to this behavior of the current assessment, and investigate potential modifications to remediate to some of the issues identified, WGWIDE recommends that an interbenchmark process should be conducted as soon as possible, preferably early in 2019. This interbenchmark would have the following tasks:

1) Improve the understanding on the behavior of the current model and its sensitivity to each data source :

- Understand the behavior and the importance of the process error in the current assessment.
- Look for any retrospective pattern in the process error
- Investigate the process error for leave one out runs (to check if the pattern is due to any particular survey)
- Exploration of model behavior with a process error variance fixed at a very low value.
- Quantify the relative weight of the different data sources. Compute a metrics that would measure the weight of each individual data point. . Individual data sources are not necessary ignored but their weight reduced/increased and the contribution to the likelihood of that and other components investigated similar what is done in the Gadget model ( Stefánsson 2003 and Taylor et al. 2007)
- Understand how the tagging data influence the model :
- Why does their exclusion result in a much larger stock
- How does each new year of recapture influence abundance at age estimates (i.e. how far back in time, which age range).

2) Investigate possible changes in the model :

- Revisit the formulation used to incorporate the tagging data. (e.g. how to interpret output of a model having at the same time normal and negative binomial error distributions)
- Revisit the selection operated on the RFID tag data (inclusion of age 2, age 12 not treated as a plus group, number of years spent before recapture)
- Consider fixing the relative weight of the tags (probably down weight them) compared to the surveys (possibly based on external information).
- post release survival estimated by periods of years (reflecting the tagging practices) instead of estimated by type of tag used. The idea is that post release survival for the last years of the steel tags should be similar to the survival for the RFID tags (since tagging protocol changed before steel tags where replaced by RFID tags).
- Revisit the down weighting of the catches prior to 2000, and consider either estimating a catch multiplier (as it is possible to do in SAM) or fix the multiplier based on external information (and test the sensitivity to the multiplier used).
- Investigate alternative use of the tagging data (e.g. as an biomass index or a Z estimate)
- Reassess the relevance of using a correlation structure for the IESSNS. Potentially use the empirical correlation in the index (estimated by stoX) with an additional variability
- Revisit other aspect of model configuration (is it overparameterised or should we give it more freedom, for instance in the observation error for the catches).


### 1.12.2 Blue Whiting

Numerous scientific studies have suggested that blue whiting in the North Atlantic consists of multiple stock units. The ICES Stock Identification Methods Working Group (SIMWG) reviewed this evidence in 2014 (ICES, SIMWG 2014) and concluded that the perception of blue whiting in the NE Atlantic as a single-stock unit is not supported by
the best available science. SIMWG further recommended that blue whiting be considered as two units. There is currently no information available that can be used as the basis for generating advice on the status of the individual stocks. However, there are some studies going on and more data being collected to allow clarify the stock definition for this species. In the future, the newly collected information on stock composition should be evaluated on the behalf of a benchmark of this stock.

Since 2016, the summer survey in the Nordic Seas (IESSNS) has provided acoustical survey indices for blue whiting (ICES. 2016c), the relevance of including that new tuning series in the assessment should be explored in next benchmark process.

### 1.12.3 NSS Herring

The 2016 benchmark assessment of Norwegian spring-spawning herring tackled most of the issues raised in last year's WGWIDE report with the aim of improving the assessment of the stock. The remaining issues and general future research of relevance for the assessment includes the following:

The Norwegian spawning ground survey was reintroduced in 2015 as part of the tuning series (fleet 1). However, changes had been made to the survey compared to the older part of the series. The 2016 benchmark accepted the inclusion of the surveys from 2015 as part of the same tuning series, but it would be relevant to explore further if the series since 2015 should be a separate tuning series due to the changes in the survey, particularly since 2019 will provide the fifth estimate from the survey since it was reintroduced.

The relevance of inclusion of a new tuning series (IESSNS) in the assessment.
Get information about uncertainty in catches from all countries (currently only available from Norway).

### 1.12.4 Western horse mackerel

Considering the potential of mixing between Western and North Sea horse mackerel occurring in Division 7.d and 7.e, better insight into the origin of catches from that area will be a major benefit for improvement of the quality of future scientific advice and thus management of the North Sea and Western horse mackerel stocks. A project addressing stock structure and boundaries of horse mackerel was initiated by the Pelagic Freezer-trawler Association (PFA) and other pelagic industries, in collaboration with Wageningen Marine Research (formerly: IMARES) and University College Dublin recently. Some preliminary results were presented to WGWIDE in 2016 but none in 2017. Further work is ongoing. The WG acknowledge the importance of this kind of research and encourage further work in this field.

Further analysis on the mixing between the Western stock and the Southern stock in area 8c should be carried out: the fishery in the area targets mainly juveniles, would be therefore be very important to understand the impact of this fishery on each of the two stocks.

### 1.12.5 North Sea horse mackerel

To improve the knowledge base for North Sea horse mackerel about the degree of connection and migrations in between the North Sea and the Western Stock, catch sampling carried out by several pelagic fishery companies is being explored to give information on the separation between North Sea and Western horse mackerel. To improve the abundance indicators the potential application of a commercial fishery
search time index will be explored. Horse mackerel is fished while it is very close to the bottom in relatively dispersed, small schools. The fishery is mostly executed using long hauls and there may be extensive search time involved. Handled in an appropriate statistical framework, taking into account the nature of the fishery and other factors such as seasonality and alternative fishing opportunities, the search time and catch rates could provide for an indication of changes in stock size over time. Catch rates in areas 27.7.e, 27.7.d and southern North Sea will be analysed from skippers' private logbooks.

The exploration of additional survey data has already been initiated in 2015 and resulted in the inclusion of the French CGFS index into the assessment of North Sea horse mackerel. In January 2017, the North Sea horse mackerel was benchmarked (ICES, 2017a). Based on capacity to model the overdispersion and the high proportion of zero values in the survey catch data the hurdle models was concluded the best option to combine the NS-IBTS and the CGFS survey information and estimating a joint annual survey index to be used for assessing the status of the stock. Future work will focus on the assessment of the importance of considering the spatial component when modelling the joint CGFS and IBTS survey index.

### 1.12.6 Boarfish

From 2017 onwards, this stock will be included on the list of stocks sampled under the data collection framework (DCMAP). This will permit sampling of commercial catch for both length and age. However, age reading is difficult and expertise is limited. An increase in the number of age readers would help develop a time-series of commercial catch-at-age which would in turn enable the development of an age-based assessment methodology. The current ALK is static and is based on a limited number of age readings.
Improvements in the survey data can be realized through a change in sampling protocol on groundfish surveys to ensure boarfish are measured to the 0.5 cm . The acoustic time-series should continue to be developed. The current survey does not contain the stock. The use of information from other acoustic surveys should also be explored.

At WGWIDE 2018, an issue list was prepared for the stock and sent to ACOM for consideration of having a benchmark assessment in 2020.

### 1.13 Decision made on next year's meeting

The WG aim for next meeting in Vigo or Santa Cruz Tenerife, Spain, in the period 27 August - 3 September 2019.

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## 2 Blue whiting (Micromesistius poutassou) in subareas 27.1-9, 12, and 14 (Northeast Atlantic)

Blue whiting (Micromesistius poutassou) is a small pelagic gadoid that is widely distributed in the eastern part of the North Atlantic. The highest concentrations are found along the edge of the continental shelf in areas west of the British Isles and on the Rockall Bank plateau where it occurs in large schools at depths ranging between 300 and 600 meters but is also present in almost all other management areas between the Barents Sea and the Strait of Gibraltar and west to the Irminger Sea. Blue whiting reaches maturity at $2-7$ years of age. Adults undertake long annual migrations from the feeding grounds to the spawning grounds. Most of the spawning takes place between March and April, along the shelf edge and banks west of the British Isles. Juveniles are abundant in many areas, with the main nursery area believed to be the Norwegian Sea. See the Stock Annex for further details on stock biology.

### 2.1 ICES advice in 2017

ICES notes that fishing mortality has increased from a historical low in 2011 to above FMSY since 2014. Spawning-stock biomass increased since 2010 and is above MSY Btrigger. Recruitment in 2017 is estimated to be low, following a period of high recruitments.

ICES advised that when the MSY approach is applied, catches in 2018 should be no more than 1387872 tonnes.

### 2.2 The fishery in 2017

The total catch in 2017 was 1558 kt . The main fisheries on blue whiting were targeting spawning and post-spawning fish (Figures 2.2.1 and 2.2.2). Most of the catches (90\%) were taken in the first two quarters of the year and the largest part of this west of the British Isles and south and east of the Faroes. Smaller quantities were taken along the coast of Spain and Portugal. The fishery in the latter half of the year was concentrated in the central Norwegian Sea. The multinational fleet currently targeting blue whiting consists of several types of vessels. The bulk of the catch is caught with large pelagic trawlers, some with capacity to process or freeze on board. The remainder is caught by RSW vessels.

### 2.3 Input to the assessment

At the Inter-Benchmark Protocol on Blue Whiting (IBPBLW 2016) it was decided to use preliminary catch-at-age data from 2016 in the assessment to get additional information to the within year IBWSS result. In most recent years more than $90 \%$ of the annual catches of the age 3+ fish are taken in the first half year, which makes it reasonable to estimate the total annual catch-at-age from reported first semester data. The catch data sections in this report give first a comprehensive description of the 2017 data as reported to ICES and then a section including a brief description of the 2018 preliminary catch data.

### 2.3.1 Officially reported catch data

Official catches in 2017 were estimated to 1558061 tonnes based on data provided by WGWIDE members. Data provided as catch by rectangle represented more than $99 \%$ of the total WG catch in 2017. Total catch by country for the period 1988 to 2017 is presented in Table 2.3.1.1 and in Figure 2.3.1.1.

After a minimum of 104000 tonnes in 2011, catches peaked in 2017 (1558 061 tonnes tonnes) (Figure 2.3.1.2.A).The spatial and temporal distribution in 2017 (Figure 2.2.1, 2.2.2 and Table 2.3.1.2), is quite similar to the distribution in previous years. The majority of catches is coming from the spawning area. The 2017 catches have largest contribution from ICES area 27.5.b, 27.6.a and 27.7.c (Figures 2.3.1.1 to 2.3.1.8). The temporal allocation of catches has been relatively stable in recent years (Figure 2.3.1.4). In the first two quarters, catches are taken over a broad area, with the highest catches in 27.5.b, 27.6.a, 27.7.c and 27.7.k, while later in the year catches is mainly taken further north in area 27.2.a and in the North Sea (27.4.a) (Figure 2.3.1.6 and 2.3.1.7 and Table 2.3.1.3). The proportion of catches originating from the Northern areas has been decreasing from 2014 to 2016, in 2017 an increase of $8 \%$ was observed.

Discards of blue whiting are small. Most of the blue whiting caught in directed fisheries are used for reduction to fish meal and fish oil. However, some discarding occurs in the fisheries for human consumption and as bycatch in fisheries directed towards other species.

Reports on discarding from fisheries which catch blue whiting were available from the Netherlands for the years 2002-2007 and 2012-2014. A study carried out to examine discarding in the Dutch fleet found that blue whiting made a minor contribution to the total pelagic discards when compared with the main species mackerel, horse mackerel and herring.

The blue whiting discards data produced by Portuguese vessels operating with bottom otter trawl within the Portuguese reaches of ICES Division 27.9.a is available since 2004. The discards data are from two fisheries: the crustacean fishery and the demersal fishery. The blue whiting estimates of discards in the crustacean fishery for the period of 2004-2011 ranged between $23 \%$ and $40 \%$ (in weight). For the same period the frequency of occurrence in the demersal fishery was around zero for the most of the years, in the years were it was significant $(2004,2006,2010)$ was ranging between $43 \%$ and $38 \%$ (in weight). In 2017, discards were $21 \%$ of the total catches for blue whiting in the Portuguese coast (Table 2.3.1.5). The total catch from Portugal is less than a half percentage of the total international catches.

Information on discards was available for Spanish fleets since 2006. Blue whiting is a bycatch in several bottom-trawl mixed fisheries. The estimates of discards in these mixed fisheries in 2006 ranged between $23 \%$ and $99 \%$ (in weight) as most of the catch is discarded and only last day catch may be retained for marketing fresh. The catch rates of blue whiting in these fisheries are however low. In the directed fishery for blue whiting for human consumption with pair trawls, discards were estimated to be 5\% (in weight) in 2015 (Table 2.3.1.5). Spanish catches are around $2 \%$ of the international catches.

In general, discards are assumed to be small in the blue whiting directed fishery. Discard data are provided by the Denmark, Portugal, Spain, UK (England and Wales) and UK(Scotland), to the working group. The discards constituted $0.13 \%$ of the total catches, 2030 tonnes.

The total estimated catches (tonnes) inside and outside the NEAFC regulatory area by country were reported on Table 2.3.1.6. Lithuania and Sweden have not provided data concerning NEAFC area, but their catches are negligible.

### 2.3.1.1 Sampling intensity

Sampling intensity for blue whiting with detailed information on the number of samples, number of fish measured, and number of fish aged by country and quarter is given in Table 2.3.1.1.1 and are presented and described by year, country and area (Tables 2.3.1.1.2 Table 2.3.1.4). In total 1779 samples were collected from the fisheries in 2017, 147297 fish were measured and 15828 were aged. The percentage of catches covered by the sampling program was $91 \%$ in 2017. The most intensive sampling took place in the area 27.4.a, 27.5.b, 27.7.c, 27.8.b, 27.8.c and 27.9.a. No sampling was carried out by Greenland, Lithuania, Poland, Sweden and the UK (England, Wales, Northern Ireland) representing together $3 \%$ of the total catches. The sampled and estimated catch-at-age data are shown on Figure 2.3.1.1.1.
Sampling intensity for age and weight of blue whiting are made in proportion to landings according to CR 1639/2001 and apply to EU member states. The Fisheries Regulation 1639/2001, requires EU Member States to take a minimum of one sample for every 1000 tonnes landed in their country. Various national sampling programs are in force.

### 2.3.1.2 Length compositions

The length distribution in numbers of around $67 \%$ of catches was provided for some of the areas sampled (Figure 2.3.1.2.1), fish from length between 20 and 30 cm dominated the catches composition.

### 2.3.1.3 Age compositions

The age-length key for the sampled catches on ICES area 27.6.a (as an example) is presented by quarter and country (Figure 2.3.1.3.1). The mean length (mm) by ages reveals that age classifications do not present significant differences between countries.

The Inter Catch program was used to calculate the total international catch-at-age, and to document how it was done.

### 2.3.2 Preliminary 2018 catch data (Quarters 1 and 2)

The preliminary catches in 2018, for quarters 1 and 2, were estimated to 1351802 tonnes (Table 2.3.2.1) based on data provided WGWIDE members.

The spatial distribution of these 2018 preliminary catches is similar to the distribution in 2017. The majority of catches are coming from the areas 27.5.b, 27.6.a, 27.6.b, 27.7.c and 27.7.k (Figure 2.3.2.1 and Table 2.3.2.2).

Sampling intensity for blue whiting from the preliminary catches by area and quarter with detailed information on the number of samples, number of fish measured, and number of fish aged is presented in Table 2.3.2.2. The preliminary catches for 2018, quarters 1 and 2 , and the expected whole 2018 catches were reported by the WGWIDE members (Table 2.3.2.3).

A comparison of the preliminary and the final catch for 2016 and 2017 (Table 2.3.2.4) shows a good agreement (i.e. max 2.9 \% deviation).

### 2.3.2.1 Raising procedure

The 2016 Benchmark concluded that the first semester(=first half year=quarter 1 and quarter 2) catch-at-ages for the preliminary year are raised to annual total catch-at-age from a 3 years average of the observed proportion of annual catches, taken in the first semester. Average proportion landed in the first semester and raising factor by age are presented in Table 2.3.2.1.1.

The WGWIDE Advice Drafting Group in 2016 proposed to further raise the preliminary first semester catches to "best available estimate" on the final catch weight. This approach is easier to communicate to the public as the raised catch is the same at the expected. The approach suggested by the ADG has been used since the 2016.
WGWIDE estimated the expected total catch for 2018 from the sum of declared national quotas, corrected for expected national uptake and transfer of these quotas (Table 2.3.2.3).

### 2.3.3 Catch-at-age

Catch-at-age numbers are presented in Table 2.3.3.1. Catch proportions at age are plotted in Figure 2.3.3.1. Strong year classes that dominated the catches can be clearly seen in the early 1980s, 1990 and the late 1990s. In recent years, the age compositions are dominated by the younger ages (ages 35).
Catch curves for the international catch-at-age dataset (Figure 2.3.3.2) indicate a consistent decline in catch number by cohort and thereby reasonably good quality catch-at-age data. Catch curves for year class 2004-2008 show a more flat curve compared to previous year classes indicating a lower $F$ or changed exploitation pattern, probably related to the low year-class strengths for some of the year classes. Year classes 20082010 show a consistent decline in the stock numbers with an estimated total mortality $(\mathrm{Z}=\mathrm{F}+\mathrm{M})$ around $0.6-0.7$ for the ages fully recruited to the fisheries.

### 2.3.4 Weight at age

Table 2.3.4.1 and Figure 2.3.4.1show the mean weight-at-age for the total catch during 1983-2018 used in the stock assessment. Mean weight at age for ages 3-9 reached a minimum around 2007, followed by an increase until 2010-2012, and a decrease in the recent years. Mean weight for the preliminary 2018 catches are calculated as the mean weight of catches in the period 2015-2017.

The weight-at-age for the stock is assumed the same as the weight-at-age for the catch.

### 2.3.5 Maturity and natural mortality

Blue whiting natural mortality and proportion of maturation-at-age are shown in Table 2.3.5.1. See the Stock Annex for further details.

### 2.3.6 Information from the fishing industry

No new information available.

### 2.3.7 Fisheries independent data

Data from the International Blue Whiting spawning stock survey are used by the stock assessment model, while recruitment indices from several other surveys are used to qualitatively adjust the most recent recruitment estimate by the assessment model and to guide the recruitments used in the forecast.

### 2.3.7.1 International Blue Whiting spawning stock survey

The Stock annex gives an overview of the surveys available for the blue whiting. The International Blue Whiting Spawning Stock Survey (IBWSS) is however the only survey used as input to the assessment model. The cruise report from IBWSS in spring 2018 is available as a working document to this report. The survey group considers that the 2018 estimate of abundance as robust.

The updated survey time-series (2004-2018) show an high internal consistency for the main age groups are given in Figure 2.3.7.1.1. B.

The distribution of acoustic backscattering densities for blue whiting for the last 4 years is shown in Figure 2.3.7.1.2. The bulk of the mature stock was located from the north Porcupine to the Hebrides core area in a corridor close to the shelf edge. This is comparable to what was observed in 2017.

The abundance estimate of blue whiting for IBWSS are presented in Table 2.3.7.1.1. In comparison to the results in 2017, there is an increase in the observed stock biomass ( $+29 \%$ ) and in stock numbers ( $+15 \%$ ).

The stock biomass within the survey area was dominated by 3, 4, 5 year old fish, contributing over $86 \%$ of total-stock biomass. The age structure of the 2018 estimate is consistent with the age structure from the 2017 estimate.

Length and age distributions for the period 2014 to 2018 are given in Figure 2.3.7.1.3.
Survey indices as applied in the stock assessment are shown in Table 2.3.7.1.2. (identical to the numbers, ages 1-8, in Table 2.3.7.1.1)

### 2.3.7.2 Other surveys

The Stock Annex provides information and time-series from surveys covering parts of the stock area. A brief survey description and survey results are provided below.

The International ecosystem survey in the Nordic Seas (IESNS) in May which is aimed at observing the pelagic ecosystem with particular focus on Norwegian spring-spawning herring and blue whiting (mainly immature fish) in the Norwegian Sea (Table 2.3.7.2.1).

Norwegian bottom-trawl survey in the Barents Sea (BS-NoRu-Q1(Btr)) in FebruaryMarch where blue whiting are regularly caught as a bycatch species. This survey gives the first reliable indication of year-class strength of blue whiting. 1 group is defined in this survey as less than 19 cm (Table 2.3.7.2.2).

Icelandic bottom-trawl surveys on the shelf and slope area around Iceland. Blue whiting is caught as bycatch species and 1-group is defined as greater than 15 cm and less than 22 cm in March (Table 2.3.7.2.3).

Faroese bottom-trawl survey on the Faroe plateau in spring where blue whiting is caught as bycatch species. 1 group is defined in this survey as less than 23 cm in March (Table 2.3.7.2.4).

The International Survey in Nordic Seas and adjacent waters in July-August (IESSNS). Blue whiting are from 2016 included as a main target species in this survey and methods are changed to sample blue whiting. This was a recommendation from WGWIDE 2015 to try to have one more time-series for blue whiting. The time-series is currently too short for assessment purposes.

This year, IEO joined the IBWSS, covering the adjacent area of the core spawning ground in Porcupine Seabight from 14th to 20th March, thus before the coverage of the core area. Blue whiting occurred in a pelagic layer located as usually at around 500 m depth, from the slope to open sea. In the southern part (from $49^{\circ} \mathrm{N}$ up to $51^{\circ} 30^{\prime} \mathrm{N}$ ) the outer limit was reached while northwards, there was a continuity towards Porcupine Bank (Figure 2.3.7.2.1). A total of 100 kt were assessed, corresponding to 1.1 billion fish. Length distribution shown 3 modes, located 19, 25 and 29 cm , with mean length esti-
mated at 25.2 cm (Figure 2.3.7.2.2). This length distribution had not significant differences with that estimated in the Spanish area (see Carrera et al. WD004 for further details). Data from the IEO survey were not included in the IBWSSS survey index.

### 2.4 Stock assessment

The presented assessment in this report follows the recommendations from the InterBenchmark Protocol of Blue Whiting (IBPBLW) convened by correspondence from 10 March to 10 May 2016 (ICES, 2016) to use the SAM model.

The configuration of the SAM model (see the Stock Annex for details) includes the same settings as agreed during IBPBLW 2016, but due to a new version of SAM, the actual values have changed in 2017. The new SAM version begins with 0 for parameters, while the old version begins with 1 . The Stock Annex has been updated accordingly.

For a model as SAM, Berg and Nielsen (2016) pointed out that the so-called "One Step Ahead" (OSA) residuals should be used for diagnostic purposes. The OSA residuals (Figure 2.4.1) show a quite random distribution of residuals. There might be an indication of "years effect" (too low index) for the IBWSS 2015 observations.

The estimated parameters from the SAM model from this year's assessment and from previous years (retrospective analysis) are shown in Table 2.4.1. There are only a very few abrupt changes in the estimated parameters over the time-series presented. The increase in process error for age 1 in the 2017 run is probably a reflection of the low 2017 recruitment. Process errors for N ages 2-10 increase slightly for the period20152018. Observation noises for the IBWSS decrease from 2017 to 2018, except for the for ages 7-8. The lowest observation noise and thereby the largest influence on the stock assessment is from catches, age 3-8, and these ages also contribute most to the international catches.

The process error residuals ("Joint sample residuals") (Figure 2.4.2) are reasonable randomly distributed.

The correlation matrix between ages for the catches and survey indices (Figure 2.4.3) show a modest observation correlation for the younger ages and stronger correlation for the older ages. The same is seen for survey observations.

Figure 2.4.4 presents estimated F at age and exploitation pattern for the whole timeseries. There are no abrupt changes in the exploitation pattern from 2010 to 2018, even though the landings in 2011 were just $19 \%$ of the landings in 2010, which might have given a different fishing practice. The estimated rather stable exploitation pattern might be due to the use of correlated random walks for $F$ at age with a high estimated correlation coefficient (rho $=0.94$, Table 2.4.2.1). However, the rather large changes in exploitation pattern for age 8 and $9+$ in the most recent years might be due to aging problems.

The retrospective analysis (Figure 2.4.5) shows an unstable assessment with substantial downward revision of SSB in the 2015 assessment(due to the 2015 low survey indices) followed by an increase in 2016. The use of "preliminary" catches (here in the retrospective analysis it is actually the final catches that are used for the period before 2017) gives a more stable assessment in the most recent 3 years. The Mohn's rho by year and as the average value over the last five years are presented in (Table 2.4.2). Even though the annual values might be high (reflecting large changes from one year to the next) the average Mohn's rho is rather low indicating no bias.

Stock summary results with added $95 \%$ confidence limits (Figure 2.4.6 and Table 2.4.5) show a decrease in fishing mortality in the period 2004-2011, followed by a steep increase in F up to 2015 and a small decrease in F in 2016-2018. Recruitment increased from low recruitments in 2006-2009 to a historically high recruitment in 2015. This is followed by a lower recruitment in 2016 and a much lower recruitment in 2017-2018. SSB has increased since 2010, followed by a small reduction in 2018.

### 2.4.1 Alternative model runs

The assessment models TISVPA and XSA were run for a better screening of potential errors in input and for comparison with the SAM results. All three models gave a similar result with respect to F, SSB and recruitment (Figure 2.4.1.1).

### 2.5 Final assessment

Following the recommendations from Inter-Benchmark Protocol on Blue Whiting (IBPBLW 2016) the SAM model is used for the final assessment. The model settings can be found in the Stock annex. Alternative model runs give similar results.

Input data are catch numbers-at-age (Table 2.3.3.1), mean weight-at-age in the stock and in the catch (Table 2.3.4.1) and natural mortality and proportion mature in Table 2.3.5.1. Applied survey data are presented in Table 2.3.7.1.2

The model was run for the period 1981-2018, with catch data up to 2017 and preliminary catch data for the first semester of 2018 raised to expected annual catches, and survey data from March-April, 2004-2018. SSB 1st January in 2019 is estimated from survivors and estimated recruits (for 2019 estimator outside the model, see short-term forecast section). $11 \%$ of age group 1 is assumed mature thus recruitment influences the size of SSB. The key results are presented in Tables 2.4.2-2.4.3 and summarized in Table 2.4.5 and Figure 2.4.6. Residuals of the model fit are shown in Figures 2.4.1-2.4.2.

### 2.6 State of the Stock

F has increased from a historic low at 0.052 in 2011 to 0.518 in 2015 followed by a decrease in $F$ to 0.454 in 2018. F has been above $\operatorname{FMSY}(0.32)$ since 2014. SSB increased from 2010 ( 2.68 million tonnes) to 2017 ( 5.50 million tonnes), followed by a decline to 2019 ( 4.33 million tonnes). SSB has been above $B_{p a}$ ( 2.25 million tonnes) since 1997.

Recruitment (age 1 fish) in 2006-2009 are in the very low end of the historical recruitments, but recruitment 2010-2016 are estimated much higher. The uncertainty around the recruitment in the most recent year is high, but recruitments in 2017 and 2018 are estimated low.

### 2.7 Biological reference points

In spring of 2016, the Inter-Benchmark Protocol on Blue Whiting (IBPBLW 2016) delegated the task of re-evaluating biological reference points of the stock to the ICES Workshop on Blue Whiting Long Term Management Strategy Evaluation (WKBWMSE). During the WGWIDE meeting 2017, WKBWMSE concluded to keep Blim and $B_{p a}$ unchanged but revised $F_{\text {lim, }}, \mathrm{F}_{\mathrm{pa}}$, and $\mathrm{F}_{\mathrm{msy}}$ (See Table below)

The table below summaries the currently used reference points.

| Framework | Reference POINT | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY Btrigger | $2.25$ <br> million <br> t | $\mathrm{B}_{\mathrm{pa}}$ | $\begin{aligned} & \text { ICES (2013a, } \\ & \text { 2013b, 2016b) } \end{aligned}$ |
|  | FMSY | 0.32 | Stochastic simulations with segmented regression stockrecruitment relationship | ICES (2016b) |
| Precautionary approach | Blim | $1.50$ <br> million t | Approximately Bloss | $\begin{aligned} & \text { ICES (2013a, } \\ & \text { 2013b, 2016b) } \end{aligned}$ |
|  | $\mathrm{B}_{\mathrm{pa}}$ | $2.25$ <br> million t | Blim $\exp (1.645 \times \sigma)$, with $\sigma=0.246$ | $\begin{aligned} & \text { ICES (2013a, } \\ & \text { 2013b, 2016b) } \end{aligned}$ |
|  | Flim | 0.88 | Equilibrium scenarios with stochastic recruitment: $F$ value corresponding to $50 \%$ probability of (SSB<Blim) | ICES (2016b) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.53 | Based on Flim and assessment uncertainties. Flim $\exp (-1.645 \times \sigma)$, with $\sigma=0.299$ | ICES (2016b) |

## References

ICES.2013a. NEAFC request to ICES to evaluate the harvest control rule element of the long-term management plan for blue whiting. Special request, Advice May 2013. In Report of the ICES Advisory Committee, 2013.ICES Advice 2013, Book 9, Section 9.3.3.1.

ICES.2013b. NEAFC request on additional management plan evaluation for blue whiting. Special request, Advice October 2013.In Report of the ICES Advisory Committee, 2013.ICES Advice 2013, Book 9, Section 9.3.3.7.

ICES. 2016b. Report of the Workshop on Blue Whiting Long Term Management Strategy Evaluation (WKBWMS), 30 August 2016ICES HQ, Copenhagen, Denmark. ICES CM 2016/ACOM:53

### 2.8 Short-term forecast

### 2.8.1 Recruitment estimates

The benchmark WKPELA in February 2012 concluded that the available survey indices should be used in a qualitative way to estimate recruitment, rather than using them in a strict quantitative model framework. The WGWIDE has followed this recommendation and investigated several survey time-series indices with the potential to give quantitative or semi-quantitative information of blue whiting recruitment. The investigated survey series were standardized by dividing with their mean and are shown in Figure 2.8.1.1.

The International Ecosystem Survey in the Nordic Seas (IESNS) only partially covers the known distribution of recruitment from this stock. Both the 1-group (2017 year class) and 2-group ( 2016 year class) indices from the survey in 2018 were below the median of the historical range.

The International Blue Whiting Spawning Stock Survey (IBWSS) is not designed to give a representative estimate of immature blue whiting. However, the 1-group indices appear to be fairly consistent with corresponding indices from older ages. The 1-group (2017 year class) index from the survey in 2018 was the below the middle of the historic range. The 2-group in 2018 (2016 year class) was in the low end in the time-series.
The Norwegian bottom-trawl survey in the Barents Sea (BS-NoRu-Q1(Btr)) in Febru-ary-March 2018, showed that 1-group blue whiting was absent (Table 2.3.7.2.2). This index should be used as a presence/absence index, in the way that when blue whiting is present in the Barents Sea, this is usually a sign of a strong year class, as all known strong year classes have been strong also in the Barents Sea

The 1-group estimate in 2018 (2017 year class) from the Icelandic bottom-trawl survey showed a decrease compared to 2017 and was in the low end in the time-series.

The 1-group estimate in 2018 (2017 year class) from the Faroese Plateau spring bottomtrawl survey was lower than in 2016 and were below the median of the historical range.

In conclusion, the indices from available survey time-series indicate that the 2016 year class is in the low end and it corresponds to the SAM assessment results. The 2017 year classes estimated from surveys are also in the low end, which also is the result of the SAM assessment where it is in the lower end. It was therefore decided not to change the SAM estimate of the 2016 and 2017 year classes.

No information is available for the 2018 and 2019 year classes and the geometric mean of the full time-series (1981-2017) was used for these year classes ( 14.6 billion at age 1 in 2018) (Table 2.8.1.1).

### 2.8.2 Short-term forecast

As decided at WGWIDE 2014, a deterministic version of the SAM forecast was applied.

### 2.8.2.1 Input

Table 2.8.2.1 lists the input data for the short-term predictions. Mean weight at age in the stock and mean weight in the catch are the same and are calculated as three year averages (2015-2017). The 2018 mean weights in the assessment are a three years average (2015-2017). Selection (exploitation pattern) is based on F in the most recent year. The proportion mature for this stock is assumed constant over the years and values are copied from the assessment input.

Recruitment (age 1) in 2017 and 2018 are assumed as estimated by the SAM model, as additional survey information was not conflicting this result. The recruitment in 2019 and 2020 are assumed at the long-term average (geometric mean for the full time-series, minus the last year (1981-2017).

As the assessment uses preliminary catches for 2018 an estimate of stock size exist for the 1 January 2019. The normal use of an "intermediate year "calculation is not relevant anymore. F in the "intermediate year" (2018) is as calculated by the assessment model. Catches in 2018 is the (model input) preliminary catches (1712874 tonnes). Intermediate year assumptions are summarised in Table 2.8.2.1.1

### 2.8.2.2 Output

A range of predicted catch and SSB options from the deterministic short-term forecast used for advice are presented in Table 2.8.2.2.1.

Following the ICES MSY framework implies fishing mortality to be at $\mathrm{F}_{\text {MSY }}=0.32$ which will give a TAC in 2018 at 1143629 tonnes ( $33.2 \%$ decrease compared to the ICES estimate of catches in 2018 and 17.2 \% decrease compared to the ICES advice for 2018). SSB is predicted to decrease by $13.2 \%$.

### 2.9 Comparison with previous assessment and forecast

Comparison of the final assessment results from the last 5 years is presented in Figure 2.9.1. The last three assessments, with the inclusion of the preliminary catches in 2016, show a tendency for overestimating SSB and underestimating F.

### 2.10 Quality considerations

Based on the confidence interval produced by the assessment model SAM there is a moderate to high uncertainty of the absolute estimate of F and SSB and the recruiting year classes (Figure 2.4.6). The retrospective analysis (Figure 2.4.5), the comparison of SSB and F estimated by three different assessment programs TISVPA, XSA and SAM (Figure 2.4.3.1) and the comparison of the 2010-2017 assessments (Figure 2.9.1) suggest a consistent assessment for the last three years (with inclusion of preliminary catch data). The preliminary 2016 and 2017 catches in weight correspond well with the final catch statistics (Table 2.3.2.4).

There are several sources of uncertainty: age reading, stock identity, and survey indices. As there is only one survey (IBWSS) that covers the spawning stock, the quality of the survey influences the assessment result considerably. The Inter-Benchmark Protocol on Blue Whiting (IBPBLW 2016) introduced a configuration of the SAM model that includes the use of estimated correlation for catch and survey observations. This handles the "year effects" in the survey observation in a better way than assuming an uncorrelated variance structure as usually applied in assessment models. However, biased survey indices will still give a biased stock estimate with the new SAM configuration.

During the WGWIDE 2017 (ICES 2017), a comparison between the mean length-at-age, by quarter and ICES division was been made. This comparison reveals a considerable lower mean length-at-age from the Faroese catch-at-age data. The 2017 catch-at-age from Faroese Islands, provided for this year assessment, were based on the age reading guidelines from the last workshop on blue whiting ageing (WKARBLUE2) and no significant deviations of the mean length-at-age have been found (Figure 2.3.1.10.). The Faroese catch-at-age data from the previous years are under revision and the assessment will be updated, when the data become available.

Utilization of preliminary catch data provides the assessment with information for the most recent year in addition to the survey information. This should give a less biased assessment as potential biased survey data in the final year are supplemented by additional catch data.

### 2.11 Management considerations

The expected catches for 2018 ( 1.712 million tonnes) are considerably higher than the ICES advice for 2018 ( 1.388 million tonnes) based on the long-term management strategy agreed by the European Union, the Faroe Islands, Iceland and Norway. This higher catch in combination with the small recruitment in 2017 and 2018 lead to the reduction in the ICES TAC advice for 2019. Without a strong recruitment in 2019 the decline in stock size and TAC will probably continue.

### 2.12 Ecosystem considerations

An extensive overview of ecosystem considerations relevant for blue whiting can be found in the stock annex.

### 2.13 Regulations and their effects

There is an agreed long-term management strategy agreed by the European Union, the Faroe Islands, Iceland and Norway. However there is no agreement between the Coastal States EU, Norway, Iceland and the Faroe Island on the share of the blue whiting TAC.

WGWIDE members estimate the total expected catch from the stock to be around 1.712 million tonnes in 2018 (close to the sum of declared quotas) whereas the TAC advice, according to the long-term management strategy was $\leq 1.388$ million tonnes.

### 2.13.1 Management plans and evaluations

A response to NEAFC request to ICES to evaluate a long-term management strategy for the fisheries on the blue whiting ICES WKBWMSE was established in the fall of 2015. The ICES Advice September 2016, "NEAFC request to ICES to evaluate a longterm management strategy for the fisheries on the blue whiting (Micromesistius poutassou) stock" concluded that:

- That the harvest control rule (HCR) proposed for the Long-Term Management Strategy (LTMS) for blue whiting, as described in the request, is precautionary given the ICES estimates of Blim ( 1.5 million t ), Bpa ( 2.25 million $\mathrm{t})$, and FMSY ( 0.32 ).
- The HCR was found to be precautionary both with and without the $20 \%$ TAC change limits above Bpa. However, the $20 \%$ TAC change limits can lead to the TAC being lowered significantly if the stock is estimated to be below Bpa, while also limiting how quickly the TAC can increase once the stock is estimated to have recovered above Bpa.
- The evaluation found that including a $10 \%$ interannual quota flexibility ('banking and borrowing') in the LTMS had an insignificant effect on the performance of the HCR.


Diagram of the requested long-term management strategy to be evaluated for blue whiting. Btrigger $=\boldsymbol{B}_{p a}$.

### 2.14 References

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### 2.15 Tables

Table 2.3.1.1.Blue whiting. ICES estimated catches (tonnes) by country for the period 1988-2017.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 18941 | 26630 | 27052 | 15538 | 34356 | 41053 | 20456 | 12439 | 52101 | 26270 | 61523 | 82935 |
| Estonia |  |  |  |  | 6156 | 1033 | 4342 | 7754 | 10982 | 5678 | 6320 |  |
| Faroes | 79831 | 75083 | 48686 | 10563 | 13436 | 16506 | 24342 | 26009 | 24671 | 28546 | 71218 | 329895 |
| France |  | 2191 |  |  |  | 1195 |  | 720 | 6442 | 12446 | 7984 | 14149 |
| Germany | 5546 | 5417 | 1699 | 349 | 1332 | 100 | 2 | 6313 | 6876 | 4724 | 17969 | 22803 |
| Iceland |  | 4977 |  |  |  |  |  | 369 | 302 | 10464 | 68681 | 501493 |
| Ireland | 4646 | 2014 |  |  | 781 |  | 3 | 222 | 1709 | 25785 | 45635 | 22580 |
| Japan |  |  |  |  | 918 | 1742 | 2574 |  |  |  |  |  |
| Latvia |  |  |  |  | 10742 | 10626 | 2582 |  |  |  |  |  |
| Lithuania |  |  |  |  |  | 2046 |  |  |  |  |  |  |
| Netherlands | 800 | 2078 | 7750 | 17369 | 11036 | 18482 | 21076 | 26775 | 17669 | 24469 | 27957 | 48303 |
| Norway | 233314 | 301342 | 310938 | 137610 | 181622 | 211489 | 229643 | 339837 | 394950 | 347311 | 560568 | 834540 |
| Poland | 10 |  |  |  |  |  |  |  |  |  |  |  |
| Portugal | 5979 | 3557 | 2864 | 2813 | 4928 | 1236 | 1350 | 2285 | 3561 | 2439 | 1900 | 2651 |
| Spain | 24847 | 30108 | 29490 | 29180 | 23794 | 31020 | 28118 | 25379 | 21538 | 27683 | 27490 | 13825 |
| S weden *** | 1229 | 3062 | 1503 | 1000 | 2058 | 2867 | 3675 | 13000 | 4000 | 4568 | 9299 | 65532 |
| UK (England + Wales)**** |  |  |  |  |  |  |  |  |  |  |  |  |
| UK (Northern Ireland) |  |  |  |  |  |  |  |  |  |  |  |  |
| UK (S cotland) | 5183 | 8056 | 6019 | 3876 | 6867 | 2284 | 4470 | 10583 | 14326 | 33398 | 92383 | 27382 |
| USSR / Russia * | 177521 | 162932 | 125609 | 151226 | 177000 | 139000 | 116781 | 107220 | 86855 | 118656 | 130042 | 355319 |
| Greenland*** |  |  |  |  |  |  |  |  |  |  |  |  |
| Unallocated |  |  |  |  |  |  |  |  |  |  |  |  |
| TOTAL | 557847 | 627447 | 561610 | 369524 | 475026 | 480679 | 459414 | 578905 | 645982 | 672437 | 1128969 | 2321406 |

* From 1992 only Russia.

Table 2.3.1.1.(continued). Blue whiting. ICES estimated catches (tonnes) by country for the period 1988-2017.

| Country | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 89500 | 41450 | 54663 | 48659 | 18134 | 248 | 140 | 165 | 340 | 2167 | 35256 | 45178 | 39395 | 60868 |
| Estonia | ** |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Faroes | 322322 | 266799 | 321013 | 317859 | 225003 | 58354 | 49979 | 16405 | 43290 | 85768 | 224700 | 282502 | 282416 | 356501 |
| France |  | 8046 | 18009 | 16638 | 11723 | 8831 | 7839 | 4337 | 9799 | 8978 | 10410 | 9659 | 10345 | 13369 |
| Germany | 15293 | 22823 | 36437 | 34404 | 25259 | 5044 | 9108 | 278 | 6239 | 11418 | 24487 | 24107 | 20025 | 45555 |
| Iceland | 379643 | 265516 | 309508 | 236538 | 159307 | 120202 | 87942 | 5887 | 63056 | 104918 | 182879 | 214870 | 186914 | 228934 |
| Ireland | 75393 | 73488 | 54910 | 31132 | 22852 | 8776 | 8324 | 1195 | 7557 | 13205 | 21466 | 24785 | 27657 | 43238 |
| Lithuania |  |  | 4635 | 9812 | 5338 |  |  |  |  |  | 4717 |  | 1129 | 5300 |
| Netherlands | 95311 | 147783 | 102711 | 79875 | 78684 | 35686 | 33762 | 4595 | 26526 | 51635 | 38524 | 56397 | 58148 | 81156 |
| Norway | 957684 | 738490 | 642451 | 539587 | 418289 | 225995 | 194317 | 20539 | 118832 | 196246 | 399520 | 489439 | 310412 | 399363 |
| Poland |  |  |  |  |  |  |  |  |  |  |  |  |  | 15889 |
| Portugal | 3937 | 5190 | 5323 | 3897 | 4220 | 2043 | 1482 | 603 | 1955 | 2056 | 2150 | 2547 | 2586 | 2046 |
| Spain | 15612 | 17643 | 15173 | 13557 | 14342 | 20637 | 12891 | 2416 | 6726 | 15274 | 32065 | 29206 | 31952 | 28920 |
| Sweden | 19083 | 2960 | 101 | 464 | 4 | 3 | 50 | 1 | 4 | 199 | 2 | 32 | 42 | 90 |
| UK (England + Wales) + | 2593 | 7356 | 10035 | 12926 | 14147 | 6176 | 2475 | 27 | 1590 | 4100 | 11 | 131 | 1374 | 3447 |
| UK (Northern Ireland) |  |  |  |  |  |  |  |  |  | 1232 | 2205 | 1119 |  |  |
| UK (Scotland) | 57028 | 104539 | 72106 | 43540 | 38150 | 173 | 5496 | 1331 | 6305 | 8166 | 24630 | 30508 | 37173 | 64724 |
| Russia | 346762 | 332226 | 329100 | 236369 | 225163 | 149650 | 112553 | 45841 | 88303 | 120674 | 152256 | 185763 | 173655 | 188449 |
| Greenland |  |  |  |  |  |  |  |  |  | 2133 |  |  |  | 20212 |
| Unallocated |  |  |  |  |  |  |  |  | 3499 |  |  |  |  |  |
| TOTAL | 2380161 | 2034309 | 1976176 | 1625255 | 1260615 | 641818 | 526357 | 103620 | 384021 | 628169 | 1155279 | 1396244 | 1183224 | 1558061 |

** Reported to the EU but not to the ICES WGNPBW (Landings of 19,467 tonnes).

+ data from 2017 updated in the 2018.

Table 2.3.1.2.Blue whiting. ICES estimated catches (tonnes) by country and area for 2017.


Table 2.3.1.3.Blue whiting. ICES estimated catches (tonnes) by quarter and area for 2017.

| Area | Quarter 1 | Quarter 2 | Quarter 3 | Quarter 4 | 2017 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27.2.a | 354 | 61145 | 43238 | 19598 |  | 124335 |
| 27.3.a | 0 | 1 | 313 | 3 |  | 317 |
| 27.3.d | 0 |  |  |  |  | 0 |
| 27.4.a | 105 | 14313 | 15789 | 14932 |  | 45138 |
| 27.4.b | 1 | 16 | 2 | 0 |  | 18 |
| 27.4.c | 0 | 0 | 0 | 0 |  | 0 |
| 27.5.a |  | 7553 |  | 4650 |  | 12203 |
| 27.5.b | 57472 | 314429 | 1436 | 83082 |  | 456419 |
| 27.6.a | 86698 | 286366 | 0 | 707 | 12 | 373783 |
| 27.6.b | 59304 | 1339 | 2 | 2 | 22 | 60668 |
| 27.7.b | 2941 | 1403 | 7 | 11 |  | 4362 |
| 27.7.c | 384291 | 4540 | 17 | 26 |  | 388873 |
| 27.7.d | 0 |  |  |  |  | 0 |
| 27.7.e | 1 | 5 | 30 | 2 |  | 38 |
| 27.7.f |  | 0 |  |  |  | 0 |
| 27.7.g |  | 11 | 0 |  |  | 11 |
| 27.7.h | 1 | 19 | 2 | 2 |  | 23 |
| 27.7.j | 15 | 186 | 44 | 305 |  | 550 |
| 27.7.k | 60500 | 1 | 1 |  |  | 60502 |
| 27.8.a | 1 | 68 | 3 | 2 |  | 73 |
| 27.8.b | 62 | 76 | 35 | 72 |  | 245 |
| 27.8.c | 6074 | 5171 | 5362 | 4843 |  | 21449 |
| 27.8.d | 0 | 1 |  | 26 |  | 27 |
| 27.9.a | 1536 | 2945 | 2508 | 2007 |  | 8997 |
| 27.14.a |  | 0 | 27 |  |  | 27 |
| 27.14.b |  |  | 0 |  |  | 0 |
| Grand total | 659354 | 699587 | 68817 | 130269 | 34 | 1558061 |

[^0]Table 2.3.1.4.Blue whiting. ICES estimated catches (tonnes) from the main fisheries 1988-2017 by area.

| Area | Norwegian <br> Sea fishery <br> (SAs1+2;Divs <br> .5.a,14a-b) | Fishery in the spawning area (SA 12.; Divs. 5.b, 6.a-b, 7.a-c) | Directedand mixed fisheries in the North Sea (SA4; Div.3.a) |  | Total <br> southern <br> areas $\left\lvert\, \begin{aligned} & (\mathrm{SAs} 8+9 ; \mathrm{Di} \\ & \text { vs.7.d-k) } \end{aligned}\right.$ | Grand total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 55829 | 426037 | 45143 | 527009 | 30838 | 557847 |
| 1989 | 42615 | 475179 | 75958 | 593752 | 33695 | 627447 |
| 1990 | 2106 | 463495 | 63192 | 528793 | 32817 | 561610 |
| 1991 | 78703 | 218946 | 39872 | 337521 | 32003 | 369524 |
| 1992 | 62312 | 318018 | 65974 | 446367 | 28722 | 475026 |
| 1993 | 43240 | 347101 | 58082 | 448423 | 32256 | 480679 |
| 1994 | 22674 | 378704 | 28563 | 429941 | 29473 | 459414 |
| 1995 | 23733 | 423504 | 104004 | 551241 | 27664 | 578905 |
| 1996 | 23447 | 478077 | 119359 | 620883 | 25099 | 645982 |
| 1997 | 62570 | 514654 | 65091 | 642315 | 30122 | 672437 |
| 1998 | 177494 | 827194 | 94881 | 1099569 | 29400 | 1128969 |
| 1999 | 179639 | 943578 | 106609 | 1229826 | 26402 | 1256228 |
| 2000 | 284666 | 989131 | 114477 | 1388274 | 24654 | 1412928 |
| 2001 | 591583 | 1045100 | 118523 | 1755206 | 24964 | 1780170 |
| 2002 | 541467 | 846602 | 145652 | 1533721 | 23071 | 1556792 |
| 2003 | 931508 | 1211621 | 158180 | 2301309 | 20097 | 2321406 |
| 2004 | 921349 | 1232534 | 138593 | 2292476 | 85093 | 2377569 |
| 2005 | 405577 | 1465735 | 128033 | 1999345 | 27608 | 2026953 |
| 2006 | 404362 | 1428208 | 105239 | 1937809 | 28331 | 1966140 |
| 2007 | 172709 | 1360882 | 61105 | 1594695 | 17634 | 1612330 |
| 2008 | 68352 | 1111292 | 36061 | 1215704 | 30761 | 1246465 |
| 2009 | 46629 | 533996 | 22387 | 603012 | 32627 | 635639 |
| 2010 | 36214 | 441521 | 17545 | 495280 | 28552 | 523832 |
| 2011 | 20599 | 72279 | 7524 | 100401 | 3191 | 103592 |
| 2012 | 24391 | 324545 | 5678.346 | 354614 | 29401.78 | 384016* |
| 2013 | 31759 | 481356 | 8749.051 | 521864 | 103973.5 | 625837** |
| 2014 | 45580 | 885483 | 28596 | 959659 | 195620 | 1155279 |
| 2015 | 150828 | 895684 | 44661 | 1091173 | 305071 | 1396244 |
| 2016 | 59744 | 905087 | 55774 | 1020604 | 162583 | 1183187 |
| 2017 | 136565 | 1284105 | 45474 | 1466144 | 91917 | 1558061 |

[^1]Table 2.3.1.5. Blue whiting. ICES estimates(tonnes) of catches, landings and discards by country for 2017.

| Country | Catches | Landings | Discards | \% discards |
| :---: | :---: | :---: | :---: | :---: |
| Denmark | 60868 | 60864 | 4 | 0.01 |
| Faroe Islands | 356501 | 356501 |  | 0.00 |
| France | 13369 | 13221 | 148 | 1.11 |
| Germany | 45555 | 45555 |  | 0.00 |
| Greenland | 20212 | 20212 |  | 0.00 |
| Iceland | 228934 | 228934 |  | 0.00 |
| Ireland | 43238 | 43238 | 0 | 0.00 |
| Lithuania | 5300 | 5300 |  | 0.00 |
| Netherlands | 81156 | 81156 |  | 0.00 |
| Norway | 399363 | 399363 |  | 0.00 |
| Poland | 15889 | 15889 |  | 0.00 |
| Portugal | 2046 | 1625 | 421 | 20.58 |
| Russia | 188449 | 188449 |  | 0.00 |
| Spain | 28920 | 27500 | 1419 | 4.91 |
| Sweden | 90 | 90 |  | 0.00 |
| UK (England) | 3447 | 3442 | 4 | 0.12 |
| UK(Scotland) | 64724 | 64690 | 34 | 0.05 |
| Total | 1558061 | 1556030 | 2030 | 0.13 |

Table 2.3.1.6. Blue whiting. ICES estimated catches (tonnes) inside and outside NEAFC area for 2017 by country.

|  | Catches inside NEAFC area | Catches outside NEAFC area | Total catches |
| :---: | :---: | :---: | :---: |
| Denmark | 3935 | 56933 | 60868 |
| Faroe Islands | 45731 | 310770 | 356501 |
| France | 230* | 13139 | 13369 |
| Germany | 41471 | 4084 | 45555 |
| Greenland | 1 | 20211 | 20212 |
| Iceland | 8127 | 220807 | 228934 |
| Ireland | 9 | 43229 | 43238 |
| Lithuania** | 0 | 5300 | 5300 |
| Netherlands | 1073 | 80082 | 81156 |
| Norway | 77714 | 321649 | 399363 |
| Poland | 0 | 15889 | 15889 |
| Portugal | 0 | 2046 | 2046 |
| Russia | 84620 | 103829 | 188449 |
| Spain | 0 | 28920 | 28920 |
| Sweden** | 0 | 90 | 90 |
| UK (England) | 108 | 3339 | 3447 |
| UK(Scotland) | 0 | 64724 | 64724 |
| Total in 2017 | 263019 | 1295042 | 1558061 |

[^2]Table 2.3.1.1.1. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme, No. of samples, No. of fish measured and No. of fish aged for 2000-2017.

| Year | Catch (tonnes) | \% catch covered by sampling programme | No. samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 1412928 | * | 1136 | 125162 | 13685 |
| 2001 | 1780170 | * | 985 | 173553 | 17995 |
| 2002 | 1556792 | * | 1037 | 116895 | 19202 |
| 2003 | 2321406 | * | 1596 | 188770 | 26207 |
| 2004 | 2377569 | * | 1774 | 181235 | 27835 |
| 2005 | 2026953 | * | 1833 | 217937 | 32184 |
| 2006 | 1966140 | * | 1715 | 190533 | 27014 |
| 2007 | 1610090 | 87 | 1399 | 167652 | 23495 |
| 2008 | 1246465 | 90 | 927 | 113749 | 21844 |
| 2009 | 635639 | 88 | 705 | 79500 | 18142 |
| 2010 | 524751 | 87 | 584 | 82851 | 16323 |
| 2011 | 103591 | 85 | 697 | 84651 | 12614 |
| 2013 | 625837 | 96 | 915 | 111079 | 14633 |
| 2014 | 1155279 | 89 | 912 | 111316 | 39738 |
| 2015 | 1396244 | 94 | 1570 | 102367 | 29821 |
| 2016 | 1183187 | 89 | 1092 | 120329 | 13793 |
| 2017 | 1558061 | 91 | 1779 | 147297 | 15828 |

Table 2.3.1.1.2. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme (catch-at-age numbers), No. of samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by country for 2017.

| Country | Catch (ton) | \% catch covered by sampling programme | No. samples | No. Measured | No. Aged | No Aged/ 1000 tonnes | No Measured/ 1000 tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 60868 | 88 | 43 | 1402 | 1402 | 23 | 23 |
| Faroe Islands | 356501 | 89 | 32 | 3159 | 1710 | 5 | 9 |
| France | 13369 | 0 | 118 | 7004 | 0 | 0 | 524 |
| Germany | 45555 | 58 | 59 | 23277 | 563 | 12 | 511 |
| Greenland | 20212 | 0 | 0 | 0 | 0 | 0 | 0 |
| Iceland | 228934 | 100 | 84 | 7545 | 2112 | 9 | 33 |
| Ireland | 43238 | 94 | 14 | 2997 | 1426 | 33 | 69 |
| Lithuania | 5300 | 0 | 0 | 0 | 0 | 0 | 0 |
| Netherlands | 81156 | 83 | 74 | 16866 | 1850 | 23 | 208 |
| Norway | 399363 | 100 | 287 | 11868 | 1120 | 3 | 30 |
| Poland | 15889 | 0 | 0 | 0 | 0 | 0 | 0 |
| Portugal | 2046 | 100 | 69 | 3725 | 1045 | 511 | 1821 |
| Russia | 188449 | 100 | 41 | 36931 | 1702 | 9 | 196 |
| Spain | 28920 | 99 | 951 | 31012 | 2526 | 87 | 1072 |
| Sweden | 90 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK (England) | 3447 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK(Scotland) | 64724 | 87 | 7 | 1511 | 372 | 6 | 23 |
| Total | 1558061 | 91 | 1779 | 147297 | 15828 | 10 | 95 |

Table 2.3.1.2.3.Blue whiting. ICES estimated catches (tonnes), No. of samples, No. of fish measured and No. of fish aged by country and quarter for 2017.

|  | Catch (tonnes) | No. samples | No. Length Measured | No. Age Samples |
| :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |
| 1 | 42554 | 39 | 1243 | 1243 |
| 2 | 18246 | 4 | 159 | 159 |
| 3 | 58 | 0 | 0 | 0 |
| 4 | 9 | O | O | - |
| Total | 60868 | 43 | 1402 | 1402 |
| Faroe Islands |  |  |  |  |
| 1 | 118066 | 16 | 1504 | 810 |
| 2 | 174329 | 13 | 1365 | 700 |
| 3 | 14778 | 2 | 190 | 100 |
| 4 | 49328 | 1 | 100 | 100 |
| Total | 356501 | 32 | 3159 | 1710 |
| France |  |  |  |  |
| 1 | 4317 | 52 | 4078 | o |
| 2 | 8084 | 57 | 2488 | o |
| 3 | 37 | 0 | 0 | o |
| 4 | 931 | 9 | 438 | 0 |
| Total | 13369 | 118 | 7004 | 0 |
| Germany |  |  |  |  |
| 1 | 11380 | 0 | 0 | o |
| 2 | 27524 | 59 | 23277 | 563 |
| 3 | 3677 | 0 | 0 | 0 |
| 4 | 2973 | O | o | O |
| Total | 45555 | 59 | 23277 | 563 |
| Greenland |  |  |  |  |
| 2 | 15260 | o | o | o |
| 3 | 29 | 0 | 0 | O |
| 4 | 4924 | - | o | O |
| Total | 20212 | 0 | 0 | 0 |
| Iceland |  |  |  |  |
| 1 | 10821 | 5 | 451 | 120 |
| 2 | 169503 | 65 | 5986 | 1642 |
| 3 | 5279 | 2 | 147 | 50 |
| 4 | 43331 | 12 | 961 | 300 |
| Total | 228934 | 84 | 7545 | 2112 |
| Ireland |  |  |  |  |
| 1 | 33829 | 13 | 2817 | 1323 |
| 2 | 9400 | 1 | 180 | 103 |
| 4 | 9 | 0 | 0 | 0 |
| Total | 43238 | 14 | 2997 | 1426 |
| Lithuania |  |  |  |  |
| 1 | 5300 | 0 | 0 | 0 |
| Total | 5300 | 0 | 0 | 0 |
| Netherlands |  |  |  |  |
| 1 | 33162 | 66 | 14854 | 1650 |
| 2 | 37306 | 8 | 2012 | 200 |
| 3 | 1286 | 0 | 0 | 0 |
| 4 | 9401 | 0 | 0 | 0 |
| Total | 81156 | 74 | 16866 | 1850 |
| Norway 1 <br>  2 <br>  3 <br>  4 |  |  |  |  |
|  | 274147 | 32 | 1368 | 647 |
|  | 98041 | 71 | 2653 | 362 |
|  | 21716 | 152 | 6768 | 90 |
|  | 5459 | 32 | 1079 | 21 |
| Total | 399363 | 287 | 11868 | 1120 |
| Poland |  |  |  |  |
| 1 | 2315 | - | 0 | o |
| 2 | 13575 | 0 | 0 | 0 |
| Total | 15889 | 0 | 0 | 0 |
| Portugal 1 <br>  2 <br>  3 |  |  |  |  |
|  | 456 | 15 | 779 | 159 |
|  | 462 | 18 | 741 | 136 |
|  | 541 | 20 | 1393 | 412 |
|  | 587 | 16 | 812 | 338 |
| Total | 2046 | 69 | 3725 | 1045 |
| Russia 1 <br>  2 <br>  3 <br>  $\mathbf{4}$ |  |  |  |  |
|  | 83298 | 13 | 11113 | 597 |
|  | 84503 | 17 | 14997 | 822 |
|  | 13800 | 5 | 5503 | 120 |
|  | 6848 | 6 | 5318 | 163 |
| Total | 188449 | 41 | 36931 | 1702 |
| Spain |  |  |  |  |
| 1 | 7231 | 257 | 8702 | 602 |
| 2 | 7803 | 270 | 9456 | 446 |
| - 3 | 7425 | 194 | 5900 | 703 |
| 4 | 6460 | 230 | 6954 | 775 |
| Total | 28920 | 951 | 31012 | 2526 |
| Sweden  <br>   <br>  1 <br>  2 <br>  3 |  |  |  |  |
|  | 0 | 0 | 0 | o |
|  | $0$ | 0 | 0 | 0 |
|  | 86 | - | 0 | - |
|  | 3 | 0 | o | 0 |
| Total ${ }^{\text {T }}$ ( ${ }^{\text {UK (England) }}$ | 90 | 0 | 0 | 0 |
|  |  |  |  |  |
| 1 | 3 | 0 | 0 | 0 |
| $2$ | 3334 | 0 | 0 | O |
| 3 | 104 | 0 | 0 | O |
| 4 |  | 0 | 0 | 0 |
| Total | 3447 | 0 | 0 | 0 |
| UK(Scotland) |  |  |  |  |
| 1 | 32474 | 5 | 1159 | 250 |
| 2 | 32216 | 2 | 352 | 122 |
| 3 | 0 | 0 | 0 | O |
| - 4 | 0 | 0 | 0 | O |
| 2017 | 34 | O | 0 | O |
| Total | 64724 | 7 | 1511 | 372 |
| Grand Total | 1558061 | 1779 | 147297 | 15828 |

* Discards data from UK(Scotland) were provided by year, due to sampling intensity.

Table 2.3.1.1.4. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme, No. of samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by ICES division for 2017.

| Division | Catch (ton) | No. samples | No. Measured | $\begin{array}{r} \text { No. } \\ \text { Aged } \end{array}$ | $\begin{array}{r} \text { No Aged/ } 1000 \\ \text { tonnes } \end{array}$ | No Measured/ 1000 tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27.2.a | 124335 | 68 | 20694 | 961 | 8 | 166 |
| 27.3.a | 317 | 26 | 442 | 0 | 0 | 1392 |
| 27.3.d | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.4.a | 45138 | 219 | 14615 | 364 | 8 | 324 |
| 27.4.b | 18 | 0 | 0 | 0 | 0 | 0 |
| 27.4.c | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.5.a | 12203 | 8 | 575 | 198 | 16 | 47 |
| 27.5.b | 456419 | 105 | 30633 | 3207 | 7 | 67 |
| 27.6.a | 373783 | 98 | 12055 | 2073 | 6 | 32 |
| 27.6.b | 60668 | 9 | 689 | 240 | 4 | 11 |
| 27.7.b | 4362 | 0 | 0 | 0 | 0 | 0 |
| 27.7.c | 388873 | 163 | 25530 | 4350 | 11 | 66 |
| 27.7.d | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.7.e | 38 | 0 | 0 | 0 | 0 | 0 |
| 27.7.f | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.7.g | 11 | 18 | 136 | 0 | 0 | 12293 |
| 27.7.h | 23 | 6 | 59 | 0 | 0 | 2514 |
| 27.7.j | 550 | 0 | 0 | 0 | 0 | 0 |
| 27.7.k | 60502 | 29 | 6982 | 864 | 14 | 115 |
| 27.8.a | 73 | 0 | 0 | 0 | 0 | 0 |
| 27.8.b | 245 | 182 | 4319 | 0 | 0 | 17652 |
| 27.8.c | 21449 | 439 | 19712 | 1335 | 62 | 919 |
| 27.8.d | 27 | 0 | 0 | 0 | 0 | 0 |
| 27.9.a | 8997 | 409 | 10856 | 2236 | 249 | 1207 |
| 27.14.a | 27 | 0 | 0 | 0 | 0 | 0 |
| 27.14.b | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL | 1558061 | 1779 | 147297 | 15828 | 385 | 36807 |

Table 2.3.2.1 .Blue whiting. ICES estimated preliminary catches (tonnes) in 2018 by quarter and area.

|  | Landings |  |  |
| :---: | :---: | :---: | :---: |
| ICES div. | Quarter 1 | Quarter 2 | Total |
| 27.2.a | 346 | 21902 | 22248 |
| 27.2.a. 1 |  | 1023 | 1023 |
| 27.2.a.2 | 20 | 4421 | 4441 |
| 27.4.a | 14 | 13105 | 13119 |
| 27.4.b |  | 7 | 7 |
| 27.5.a |  | 2403 | 2403 |
| 27.5.b | 74661 | 376864 | 451525 |
| 27.5.b.1 |  | 1222 | 1222 |
| 27.6.a | 31506 | 310952 | 342457 |
| 27.6.b | 140727 | 300 | 141027 |
| 27.6.b.2 | 22083 | 788 | 22870 |
| 27.7.b | 4017 | 756 | 4773 |
| 27.7.c | 181026 | 6416 | 187442 |
| 27.7.c. 2 | 91345 | 4781 | 96127 |
| 27.7.j. 2 | 1 | 21 | 22 |
| 27.7.k | 39508 |  | 39508 |
| 27.9.a | 177 | 364 | 541 |
| Total | 585430 | 745323 | 1330754 |

Table 2.3.2.2.Blue whiting. ICES estimated preliminary catches (tonnes), the percentage of catch covered by the sampling programme, No. of samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by ICES division for 2018 preliminary data (quarters 1 and 2).

| ICES div. | Catch (ton) | No. samples | No. Measured | No. Aged |
| :--- | ---: | ---: | ---: | ---: |
| 27.2.a | 22248 | 3 | 261 | 74 |
| 27.2.a.1 | 1023 | 0 | 0 | 0 |
| 27.2.a.2 | 4441 | 0 | 0 | 0 |
| 27.4.a | 13119 | 0 | 0 | 0 |
| 27.4.b | 7 | 0 | 0 | 0 |
| 27.5.a | 2403 | 0 | 0 | 0 |
| 27.5.b | 451525 | 56 | 16213 | 2215 |
| 27.5.b.1 | 1222 | 2 | 34 | 34 |
| 27.6.a | 342457 | 22 | 4157 | 1210 |
| 27.6.b | 141027 | 14 | 2582 | 452 |
| 27.6.b.2 | 22870 | 5 | 502 | 301 |
| 27.7.b | 4773 | 0 | 0 | 0 |
| 27.7.c | 187442 | 33 | 7108 | 1338 |
| 27.7.c.2 | 96127 | 19 | 1960 | 755 |
| 27.7.j.2 | 22 | 0 | 0 | 0 |
| 27.7.k | 39508 | 26 | 5496 | 360 |
| 27.9.a | 541 | 11 | 749 | 251 |
| Total | $\mathbf{1 3 3 0 7 5 4}$ | $\mathbf{1 9 1}$ | $\mathbf{3 9 0 6 2}$ | $\mathbf{6 9 9 0}$ |

Table 2.3.2.3. Blue whiting. ICES estimates of catches (tonnes) in 2018, based on (initial) declared quotas and expected uptake estimated by WGWIDE.

| Country | Prelim Q1-Q2 Catch | National Quota | DEVIATION FROM QUOTA |
| :---: | :---: | :---: | :---: |
| Denmark | 87289 | 61,277 | 26,012 |
| Faroe Islands | 307,353 | 493,081 | -99,500 |
| Germany | 39,281 | 23,825 | 15,456 |
| Greenland | 14,839 | 16,000 | 0 |
| Iceland | 224,009 | 275971 | 0 |
| Ireland | 47,620 | 47,451 | 0 |
| Lithuania | 0 | 5,300 | 0 |
| Netherlands | 0 | 74,720 | 40,013 |
| Norway | 420,161 | 421,100 | 10,000 |
| Portugal | 541 | 4,826 | -2,413 |
| Russia | 145,206 | 207,345 | -47,345 |
| UK(Scotland) | 65,503 | 79,513 | -9,513 |
| UK (England) | 0 | 0 | 0 |
| Sweden | 0 | 15,158 | -15,000 |
| France | 0 | 42,644 | -25,000 |
| Spain | 0 | 51,949 | 0 |
| Total | 1,351,802 | 1,819,860 | -106,990 |
| EU | 240,234 | 401,363 |  |
| Best guess on catches in 2018 |  |  | 1,712,870 |

Table 2.3.2.4. Blue whiting. Comparison of preliminary and final catches (tonnes).

| YEAR | PRELIMINARY | FINAL | DEVIATION \%* |
| :---: | ---: | :---: | :---: |
| 2016 | 1147000 | 1180786 | 2.9 |
| 2017 | 1559437 | 1555069 | -0.3 |
| 2018 | 1712874 |  |  |

[^3]Table 2.3.2.1.1 .Blue whiting. Proportion of the annual catch taken in the first half-year of 2015-2017, average proportion and scaling factor used for raising the preliminary first half year of 2018 catch data.

| VaLues | 2015 | 2016 | 2017 | AVERAGE | Raising FACTOR |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Age 1 | 76.6 | 76.4 | 73.3 | 75.4 | 1.326 |
| Age 2 | 83.7 | 85.9 | 82.5 | 84.0 | 1.190 |
| Age 3 | 87.4 | 92.2 | 87.9 | 89.2 | 1.122 |
| Age 4 | 89.5 | 92.3 | 91.0 | 90.9 | 1.100 |
| Age 5 | 91.7 | 97.0 | 93.8 | 94.2 | 1.062 |
| Age 6 | 88.9 | 97.1 | 94.5 | 93.5 | 1.069 |
| Age 7 | 88.9 | 96.2 | 98.1 | 94.4 | 1.059 |
| Age 8 | 90.8 | 98.1 | 97.2 | 95.4 | 1.048 |
| Age 9 | 95.2 | 96.3 | 98.6 | 96.7 | 1.034 |
| Age 10 | 90.3 | 95.0 | 97.2 | 94.2 | 1.062 |
|  |  |  |  |  |  |

Table 2.3.3.1.Bluewhiting. Catch-at-age numbers (thousands) by year. Discards included since 2014. Values for 2018 are preliminary.

| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 258000 | 348000 | 681000 | 334000 | 548000 | 559000 | 466000 | 634000 | 578000 | 1460000 |
| 1982 | 148000 | 274000 | 326000 | 548000 | 264000 | 276000 | 266000 | 272000 | 284000 | 673000 |
| 1983 | 2283000 | 567000 | 270000 | 286000 | 299000 | 304000 | 287000 | 286000 | 225000 | 334000 |
| 1984 | 2291000 | 2331000 | 455000 | 260000 | 285000 | 445000 | 262000 | 193000 | 154000 | 255000 |
| 1985 | 1305000 | 2044000 | 1933000 | 303000 | 188000 | 321000 | 257000 | 174000 | 93000 | 259000 |
| 1986 | 650000 | 816000 | 1862000 | 1717000 | 393000 | 187000 | 201000 | 198000 | 174000 | 398000 |
| 1987 | 838000 | 578000 | 728000 | 1897000 | 726000 | 137000 | 105000 | 123000 | 103000 | 195000 |
| 1988 | 425000 | 721000 | 614000 | 683000 | 1303000 | 618000 | 84000 | 53000 | 33000 | 50000 |
| 1989 | 865000 | 718000 | 1340000 | 791000 | 837000 | 708000 | 139000 | 50000 | 25000 | 38000 |
| 1990 | 1611000 | 703000 | 672000 | 753000 | 520000 | 577000 | 299000 | 78000 | 27000 | 95000 |
| 1991 | 266686 | 1024468 | 513959 | 301627 | 363204 | 258038 | 159153 | 49431 | 5060 | 9570 |
| 1992 | 407730 | 653838 | 1641714 | 569094 | 217386 | 154044 | 109580 | 79663 | 31987 | 11706 |
| 1993 | 263184 | 305180 | 621085 | 1571236 | 411367 | 191241 | 107005 | 64769 | 38118 | 17476 |
| 1994 | 306951 | 107935 | 367962 | 389264 | 1221919 | 281120 | 174256 | 90429 | 79014 | 30614 |
| 1995 | 296100 | 353949 | 421560 | 465358 | 615994 | 800201 | 253818 | 159797 | 59670 | 41811 |
| 1996 | 1893453 | 534221 | 632361 | 537280 | 323324 | 497458 | 663133 | 232420 | 98415 | 82521 |
| 1997 | 213149 | 1519327 | 904074 | 577676 | 295671 | 251642 | 282056 | 406910 | 104320 | 169235 |
| 1998 | 1656926 | 4181175 | 3541231 | 1044897 | 383658 | 322777 | 303058 | 264105 | 212452 | 85513 |
| 1999 | 788200 | 1549100 | 5820800 | 3460600 | 412800 | 207200 | 151200 | 153100 | 68800 | 140500 |
| 2000 | 1814851 | 1192657 | 3465739 | 5014862 | 1550063 | 513663 | 213057 | 151429 | 58277 | 139791 |
| 2001 | 4363690 | 4486315 | 2962163 | 3806520 | 2592933 | 585666 | 170020 | 97032 | 76624 | 66410 |
| 2002 | 1821053 | 3232244 | 3291844 | 2242722 | 1824047 | 1647122 | 344403 | 168848 | 102576 | 142743 |
| 2003 | 3742841 | 4073497 | 8378955 | 4824590 | 2035096 | 1117179 | 400022 | 121280 | 19701 | 27493 |
| 2004 | 2156261 | 4426323 | 6723748 | 6697923 | 3044943 | 1276412 | 649885 | 249097 | 75415 | 36805 |
| 2005 | 1427277 | 1518938 | 5083550 | 5871414 | 4450171 | 1419089 | 518304 | 249443 | 100374 | 55226 |
| 2006 | 412961 | 939865 | 4206005 | 6150696 | 3833536 | 1718775 | 506198 | 181181 | 67573 | 36688 |
| 2007 | 167027 | 306898 | 1795021 | 4210891 | 3867367 | 2353478 | 935541 | 320529 | 130202 | 88573 |
| 2008 | 408790 | 179211 | 545429 | 2917190 | 3262956 | 1919264 | 736051 | 315671 | 113086 | 126637 |
| 2009 | 61125 | 156156 | 231958 | 594624 | 1596095 | 1156999 | 592090 | 251529 | 88615 | 48908 |
| 2010 | 349637 | 222975 | 160101 | 208279 | 646380 | 992214 | 702569 | 256604 | 70487 | 43693 |
| 2011 | 162997 | 101810 | 63954 | 53863 | 69717 | 116396 | 120359 | 55470 | 25943 | 12542 |
| 2012 | 239667 | 351845 | 663155 | 141854 | 106883 | 203419 | 363779 | 356785 | 212492 | 157947 |
| 2013 | 228175 | 508122 | 848597 | 896966 | 462714 | 224066 | 321310 | 397536 | 344285 | 383601 |
| 2014 | 588717 | 584084 | 2312953 | 2019373 | 1272862 | 416523 | 386396 | 462339 | 526141 | 662747 |


| Year Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 2944849 | 2852384 | 2427329 | 2465286 | 1518235 | 707533 | 329882 | 258743 | 239164 | 450046 |
| 2016 | 1239331 | 3518677 | 2933271 | 1874011 | 1367844 | 756824 | 339851 | 185368 | 131039 | 288635 |
| 2017 | 401947 | 1999011 | 7864694 | 4063916 | 1509651 | 777185 | 263007 | 110351 | 63945 | 149369 |
| 2018 | 497019 | 575187 | 3292297 | 6825720 | 3034801 | 1026145 | 312013 | 112844 | 69289 | 166324 |

Table 2.3.4.1. Blue whiting. Individual mean weight (kg) at age in the catch. Preliminary values for 2018 (average of 2015-2017) are included.

| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.052 | 0.065 | 0.103 | 0.125 | 0.141 | 0.155 | 0.170 | 0.178 | 0.187 | 0.213 |
| 1982 | 0.045 | 0.072 | 0.111 | 0.143 | 0.156 | 0.177 | 0.195 | 0.200 | 0.204 | 0.231 |
| 1983 | 0.046 | 0.074 | 0.118 | 0.140 | 0.153 | 0.176 | 0.195 | 0.200 | 0.204 | 0.228 |
| 1984 | 0.035 | 0.078 | 0.089 | 0.132 | 0.153 | 0.161 | 0.175 | 0.189 | 0.186 | 0.206 |
| 1985 | 0.038 | 0.074 | 0.097 | 0.114 | 0.157 | 0.177 | 0.199 | 0.208 | 0.218 | 0.237 |
| 1986 | 0.040 | 0.073 | 0.108 | 0.130 | 0.165 | 0.199 | 0.209 | 0.243 | 0.246 | 0.257 |
| 1987 | 0.048 | 0.086 | 0.106 | 0.124 | 0.147 | 0.177 | 0.208 | 0.221 | 0.222 | 0.254 |
| 1988 | 0.053 | 0.076 | 0.097 | 0.128 | 0.142 | 0.157 | 0.179 | 0.199 | 0.222 | 0.260 |
| 1989 | 0.059 | 0.079 | 0.103 | 0.126 | 0.148 | 0.158 | 0.171 | 0.203 | 0.224 | 0.253 |
| 1990 | 0.045 | 0.070 | 0.106 | 0.123 | 0.147 | 0.168 | 0.175 | 0.214 | 0.217 | 0.256 |
| 1991 | 0.055 | 0.091 | 0.107 | 0.136 | 0.174 | 0.190 | 0.206 | 0.230 | 0.232 | 0.266 |
| 1992 | 0.057 | 0.083 | 0.119 | 0.140 | 0.167 | 0.193 | 0.226 | 0.235 | 0.284 | 0.294 |
| 1993 | 0.066 | 0.082 | 0.109 | 0.137 | 0.163 | 0.177 | 0.200 | 0.217 | 0.225 | 0.281 |
| 1994 | 0.061 | 0.087 | 0.108 | 0.137 | 0.164 | 0.189 | 0.207 | 0.217 | 0.247 | 0.254 |
| 1995 | 0.064 | 0.091 | 0.118 | 0.143 | 0.154 | 0.167 | 0.203 | 0.206 | 0.236 | 0.256 |
| 1996 | 0.041 | 0.080 | 0.102 | 0.116 | 0.147 | 0.170 | 0.214 | 0.230 | 0.238 | 0.279 |
| 1997 | 0.047 | 0.072 | 0.102 | 0.121 | 0.140 | 0.166 | 0.177 | 0.183 | 0.203 | 0.232 |
| 1998 | 0.048 | 0.072 | 0.094 | 0.125 | 0.149 | 0.178 | 0.183 | 0.188 | 0.221 | 0.248 |
| 1999 | 0.063 | 0.078 | 0.088 | 0.109 | 0.142 | 0.170 | 0.199 | 0.193 | 0.192 | 0.245 |
| 2000 | 0.057 | 0.075 | 0.086 | 0.104 | 0.133 | 0.156 | 0.179 | 0.187 | 0.232 | 0.241 |
| 2001 | 0.050 | 0.078 | 0.094 | 0.108 | 0.129 | 0.163 | 0.186 | 0.193 | 0.231 | 0.243 |
| 2002 | 0.054 | 0.074 | 0.093 | 0.115 | 0.132 | 0.155 | 0.173 | 0.233 | 0.224 | 0.262 |
| 2003 | 0.049 | 0.075 | 0.098 | 0.108 | 0.131 | 0.148 | 0.168 | 0.193 | 0.232 | 0.258 |
| 2004 | 0.042 | 0.066 | 0.089 | 0.102 | 0.123 | 0.146 | 0.160 | 0.173 | 0.209 | 0.347 |
| 2005 | 0.039 | 0.068 | 0.084 | 0.099 | 0.113 | 0.137 | 0.156 | 0.166 | 0.195 | 0.217 |
| 2006 | 0.049 | 0.072 | 0.089 | 0.105 | 0.122 | 0.138 | 0.163 | 0.190 | 0.212 | 0.328 |
| 2007 | 0.050 | 0.064 | 0.091 | 0.103 | 0.115 | 0.130 | 0.146 | 0.169 | 0.182 | 0.249 |
| 2008 | 0.055 | 0.075 | 0.100 | 0.106 | 0.120 | 0.133 | 0.146 | 0.160 | 0.193 | 0.209 |
| 2009 | 0.056 | 0.085 | 0.105 | 0.119 | 0.124 | 0.138 | 0.149 | 0.179 | 0.214 | 0.251 |
| 2010 | 0.052 | 0.064 | 0.110 | 0.154 | 0.154 | 0.163 | 0.175 | 0.187 | 0.200 | 0.272 |
| 2011 | 0.055 | 0.079 | 0.107 | 0.136 | 0.169 | 0.169 | 0.179 | 0.189 | 0.214 | 0.270 |
| 2012 | 0.041 | 0.072 | 0.098 | 0.140 | 0.158 | 0.172 | 0.180 | 0.185 | 0.189 | 0.203 |
| 2013 | 0.051 | 0.077 | 0.094 | 0.117 | 0.139 | 0.162 | 0.185 | 0.188 | 0.198 | 0.197 |


| Year AGe | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | 0.049 | 0.078 | 0.093 | 0.112 | 0.128 | 0.155 | 0.178 | 0.190 | 0.202 | 0.217 |
| 2015 | 0.039 | 0.070 | 0.094 | 0.117 | 0.137 | 0.155 | 0.174 | 0.183 | 0.193 | 0.201 |
| 2016 | 0.047 | 0.066 | 0.084 | 0.107 | 0.125 | 0.142 | 0.152 | 0.167 | 0.184 | 0.206 |
| 2017 | 0.056 | 0.072 | 0.080 | 0.094 | 0.113 | 0.131 | 0.148 | 0.172 | 0.190 | 0.212 |
| 2018 | 0.047 | 0.069 | 0.086 | 0.106 | 0.125 | 0.143 | 0.158 | 0.174 | 0.189 | 0.206 |

Table 2.3.5.1.Blue whiting. Natural mortality and proportion mature.

| AGE | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 - 1 0 +}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Proportion <br> mature | 0.00 | 0.11 | 0.40 | 0.82 | 0.86 | 0.91 | 0.94 | 1.00 |
| Natural <br> mortality | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |

Table 2.3.7.1.1.Bluewhiting.Time-series of StoX abundance estimates of blue whiting (millions) by age in the IBWSS. Total biomass in last column ( 1000 t ).

| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ | TSB |  |  |  |  |  |  |  |  |
| 2004 | 1097 | 5538 | 13062 | 15134 | 5119 | 1086 | 994 | 593 | 164 |  | 3505 |  |  |  |  |  |  |  |  |
| 2005 | 2129 | 1413 | 5601 | 7780 | 8500 | 2925 | 632 | 280 | 129 | 23 | 2513 |  |  |  |  |  |  |  |  |
| 2006 | 2512 | 2222 | 10858 | 11677 | 4713 | 2717 | 923 | 352 | 198 | 31 | 3512 |  |  |  |  |  |  |  |  |
| 2007 | 468 | 706 | 5241 | 11244 | 8437 | 3155 | 1110 | 456 | 123 | 58 | 3274 |  |  |  |  |  |  |  |  |
| 2008 | 337 | 523 | 1451 | 6642 | 6722 | 3869 | 1715 | 1028 | 269 | 284 | 2639 |  |  |  |  |  |  |  |  |
| 2009 | 275 | 329 | 360 | 1292 | 3739 | 3457 | 1636 | 587 | 250 | 162 | 1599 |  |  |  |  |  |  |  |  |
| $2010^{*}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2011 | 312 | 1361 | 1135 | 930 | 1043 | 1712 | 2170 | 2422 | 1298 | 250 | 1826 |  |  |  |  |  |  |  |  |
| 2012 | 1141 | 1818 | 6464 | 1022 | 596 | 1420 | 2231 | 1785 | 1256 | 1022 | 2355 |  |  |  |  |  |  |  |  |
| 2013 | 586 | 1346 | 6183 | 7197 | 2933 | 1280 | 1306 | 1396 | 927 | 1670 | 3107 |  |  |  |  |  |  |  |  |
| 2014 | 4183 | 1491 | 5239 | 8420 | 10202 | 2754 | 772 | 577 | 899 | 1585 | 3337 |  |  |  |  |  |  |  |  |
| 2015 | 3255 | 4565 | 1888 | 3630 | 1792 | 465 | 173 | 108 | 206 | 247 | 1403 |  |  |  |  |  |  |  |  |
| 2016 | 2745 | 7893 | 10164 | 6274 | 4687 | 1539 | 413 | 133 | 235 | 256 | 2873 |  |  |  |  |  |  |  |  |
| 2017 | 275 | 2180 | 15939 | 10196 | 3621 | 1711 | 900 | 75 | 66 | 144 | 3135 |  |  |  |  |  |  |  |  |
| 2018 | 836 | 628 | 6615 | 21490 | 7692 | 2187 | 755 | 188 | 72 | 144 | 4035 |  |  |  |  |  |  |  |  |

*Survey discarded.

Table 2.3.7.1.2.Blue Whiting. Survey indices (IBWSS) as used in the assessment.

| Year/ |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 |
| 2004 | 1097 | 5538 | 13062 | 15134 | 5119 | 1086 | 994 | 593 |
|  | 2005 | 2129 | 1413 | 5601 | 7780 | 8500 | 2925 | 632 |
| 2006 | 2512 | 2222 | 10858 | 11677 | 4713 | 2717 | 923 | 352 |
| 2007 | 468 | 706 | 5241 | 11244 | 8437 | 3155 | 1110 | 456 |
| 2008 | 337 | 523 | 1451 | 6642 | 6722 | 3869 | 1715 | 1028 |
| 2009 | 275 | 329 | 360 | 1292 | 3739 | 3457 | 1636 | 587 |
| 2010 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 2011 | 312 | 1361 | 1135 | 930 | 1043 | 1712 | 2170 | 2422 |
| 2012 | 1141 | 1818 | 6464 | 1022 | 596 | 1420 | 2231 | 1785 |
| 2013 | 586 | 1346 | 6183 | 7197 | 2933 | 1280 | 1306 | 1396 |
| 2014 | 4183 | 1491 | 5239 | 8420 | 10202 | 2754 | 772 | 577 |
| 2015 | 3255 | 4565 | 1888 | 3630 | 1792 | 465 | 173 | 108 |
| 2016 | 2745 | 7893 | 10164 | 6274 | 4687 | 1539 | 413 | 133 |
| 2017 | 275 | 20180 | 15939 | 10196 | 3621 | 1711 | 900 | 75 |
| 2018 | 836 | 628 | 6615 | 21490 | 7692 | 2187 | 755 | 188 |

Table 2.3.7.2.1.Blue Whiting. Estimated abundance of 1 and 2 year old blue whiting from the International Norwegian Sea ecosystem survey, 2003-2018.

| YEAR $\backslash$ AGE | AGE 1 | AGE 2 |
| :---: | :---: | :---: |
| $2003^{*}$ | 16127 | 9317 |
| $2004^{*}$ | 17792 | 11020 |
| $2005^{*}$ | 19933 | 7908 |
| $2006^{*}$ | 2512 | 5504 |
| $2007^{*}$ | 592 | 213 |
| 2008 | 25 | 17 |
| 2009 | 7 | 8 |
| 2010 | 0 | 280 |
| 2011 | 1613 | 0 |
| 2012 | 9476 | 3265 |
| 2013 | 454 | 6544 |
| 2014 | 3893 | 2048 |
| 2015 | 8563 | 2796 |
| 2016 | 4223 | 8089 |
| 2017 | 1236 | 2087 |
| 2018 | 441 | 1491 |

*Using the old TS-value. To compare the results all values were divided by approximately 3.1.

Table 2.3.7.2.2.Blue whiting.1-group indices of blue whiting from the Norwegian winter survey (late January-early March) in the Barents Sea. (Blue whiting < 19 cm in total body length which most likely belong to 1-group.)

| Catch Rate |  |  |
| :---: | :---: | :---: |
| Year | ALL | < 19 cm |
| 1981 | 0.13 | 0 |
| 1982 | 0.17 | 0.01 |
| 1983 | 4.46 | 0.46 |
| 1984 | 6.97 | 2.47 |
| 1985 | 32.51 | 0.77 |
| 1986 | 17.51 | 0.89 |
| 1987 | 8.32 | 0.02 |
| 1988 | 6.38 | 0.97 |
| 1989 | 1.65 | 0.18 |
| 1990 | 17.81 | 16.37 |
| 1991 | 48.87 | 2.11 |
| 1992 | 30.05 | 0.06 |
| 1993 | 5.80 | 0.01 |
| 1994 | 3.02 | 0 |
| 1995 | 1.65 | 0.10 |
| 1996 | 9.88 | 5.81 |
| 1997 | 187.24 | 175.26 |
| 1998 | 7.14 | 0.21 |
| 1999 | 5.98 | 0.71 |
| 2000 | 129.23 | 120.90 |
| 2001 | 329.04 | 233.76 |
| 2002 | 102.63 | 9.69 |
| 2003 | 75.25 | 15.15 |
| 2004 | 124.01 | 36.74 |
| 2005 | 206.18 | 90.23 |
| 2006 | 269.2 | 3.52 |
| 2007 | 80.38 | 0.16 |
| 2008 | 17.97 | 0.04 |
| 2009 | 4.50 | 0.01 |
| 2010 | 3.30 | 0.08 |
| 2011 | 1.48 | 0.01 |
| 2012 | 127.71 | 125.93 |
| 2013 | 39.54 | 2.33 |
| 2014 | 31.48 | 24.97 |
| 2015 | 148.4 | 128.34 |
| 2016 | 86.99 | 11.31 |
| 2017 | 167.16 | 0.71 |
| 2018 | 9.52 | 0.007 |

Table 2.3.7.2.3.Blue whiting.1-group indices of blue whiting from the Icelandic bottom-trawl surveys, 1-group (< 22 cm in March).

| Catch Rate |  |
| :---: | :---: |
| Year | $<22$ CM |
| 1996 | 6.5 |
| 1997 | 3.4 |
| 1998 | 1.1 |
| 1999 | 6.3 |
| 2000 | 9 |
| 2001 | 5.2 |
| 2002 | 14.2 |
| 2003 | 15.4 |
| 2004 | 8.9 |
| 2005 | 8.3 |
| 2006 | 30.4 |
| 2007 | 3.9 |
| 2008 | 0.1 |
| 2009 | 1.6 |
| 2010 | 0.2 |
| 2011 | 10.8 |
| 2012 | 29.9 |
| 2013 | 11.7 |
| 2014 | 66.3 |
| 2015 | 43.8 |
| 2016 | 6.3 |
| 2017 | 1.8 |
| 2018 | 0.4 |

Table 2.3.7.2.4.Blue whiting.1-group indices of blue whiting from Faroese bottom-trawl surveys, 1group ( $<23 \mathrm{~cm}$ in March).

| Catch Rate |  |
| :---: | :---: |
| Year | $<23 \mathrm{CM}$ |
| 1994 | 1382 |
| 1995 | 1105 |
| 1996 | 4442 |
| 1997 | 1764 |
| 1998 | 360 |
| 1999 | 1330 |
| 2000 | 782 |
| 2001 | 3357 |
| 2002 | 3885 |
| 2003 | 929 |
| 2004 | 15163 |
| 2005 | 23750 |
| 2006 | 13364 |
| 2007 | 11509 |
| 2008 | 840 |
| 2009 | 3754 |
| 2010 | 824 |
| 2011 | 11406 |
| 2012 | 5345 |
| 2013 | 8855 |
| 2014 | 51313 |
| 2015 | 14444 |
| 2016 | 22485 |
| 2017 | 5286 |
| 2018 | 1948 |

Table 2.4.1.Blue whiting. Parameter estimates, from final assessment (2018) and retrospective analysis (2015-2017).

| Parameter Year | 2014 | 2015 | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Random walk variance |  |  |  |  |  |
| -F Age 1-10 | 0.40 | 0.41 | 0.39 | 0.38 | 0.38 |
| Process error |  |  |  |  |  |
| -log(N) Age 1 | 0.58 | 0.58 | 0.58 | 0.62 | 0.62 |
| --- Age 2-10 | 0.15 | 0.17 | 0.17 | 0.18 | 0.19 |
| Observation variance |  |  |  |  |  |
| -Catch Age 1 | 0.41 | 0.46 | 0.45 | 0.44 | 0.44 |
| --- Age 2 | 0.30 | 0.29 | 0.29 | 0.29 | 0.28 |
| --- Age 3-8 | 0.21 | 0.20 | 0.20 | 0.20 | 0.19 |
| --- Age 9-10 | 0.41 | 0.40 | 0.40 | 0.40 | 0.40 |
| -IBWSS Age 1 | 0.91 | 0.77 | 0.75 | 0.77 | 0.73 |
| --- Age 2 | 0.33 | 0.33 | 0.31 | 0.32 | 0.31 |
| --- Age 3 | 0.42 | 0.46 | 0.46 | 0.44 | 0.42 |
| --- Age 4-6 | 0.35 | 0.45 | 0.45 | 0.40 | 0.39 |
| --- Age 7-8 | 0.29 | 0.37 | 0.41 | 0.48 | 0.51 |
| Survey_catchability |  |  |  |  |  |
| -IBWSS Age 1 | 0.06 | 0.07 | 0.07 | 0.07 | 0.07 |
| --- Age 2 | 0.10 | 0.12 | 0.12 | 0.12 | 0.12 |
| --- Age 3 | 0.33 | 0.38 | 0.36 | 0.38 | 0.38 |
| --- Age 4 | 0.60 | 0.70 | 0.66 | 0.70 | 0.70 |
| --- Age 5-8 | 0.86 | 0.92 | 0.86 | 0.89 | 0.88 |
| Rho |  |  |  |  |  |
| -- | 0.91 | 0.92 | 0.92 | 0.93 | 0.94 |

Table 2.4.2.Blue whiting. Mohn's rho by year and average over the last five years ( $n=5$ ).

| Year | R(AGE 1) | SSB | Fbar(3-7) |
| :---: | :---: | :---: | :---: |
| 2013 | -0.218 | 0.206 | -0.140 |
| 2014 | -0.350 | 0.353 | -0.298 |
| 2015 | -0.266 | -0.094 | 0.203 |
| 2016 | 0.472 | 0.134 | -0.167 |
| 2017 | 0.177 | -0.009 | 0.026 |
| Rho-mean | -0.037 | 0.118 | -0.075 |

Table 2.4.3.Blue whiting. Estimated fishing mortalities. Catch data for 2018 are preliminary.

| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.078 | 0.118 | 0.172 | 0.212 | 0.245 | 0.318 | 0.346 | 0.442 | 0.482 | 0.482 |
| 1982 | 0.067 | 0.102 | 0.148 | 0.183 | 0.208 | 0.270 | 0.293 | 0.371 | 0.401 | 0.401 |
| 1983 | 0.078 | 0.118 | 0.171 | 0.211 | 0.240 | 0.314 | 0.338 | 0.419 | 0.445 | 0.445 |
| 1984 | 0.096 | 0.143 | 0.212 | 0.265 | 0.305 | 0.397 | 0.418 | 0.509 | 0.528 | 0.528 |
| 1985 | 0.101 | 0.151 | 0.230 | 0.295 | 0.347 | 0.448 | 0.466 | 0.561 | 0.575 | 0.575 |
| 1986 | 0.113 | 0.169 | 0.269 | 0.358 | 0.432 | 0.553 | 0.573 | 0.692 | 0.703 | 0.703 |
| 1987 | 0.101 | 0.150 | 0.248 | 0.338 | 0.416 | 0.538 | 0.560 | 0.674 | 0.674 | 0.674 |
| 1988 | 0.098 | 0.148 | 0.253 | 0.349 | 0.438 | 0.574 | 0.588 | 0.693 | 0.677 | 0.677 |
| 1989 | 0.113 | 0.171 | 0.303 | 0.419 | 0.525 | 0.684 | 0.711 | 0.840 | 0.804 | 0.804 |
| 1990 | 0.105 | 0.159 | 0.292 | 0.407 | 0.510 | 0.663 | 0.711 | 0.846 | 0.813 | 0.813 |
| 1991 | 0.059 | 0.089 | 0.167 | 0.234 | 0.288 | 0.366 | 0.393 | 0.463 | 0.448 | 0.448 |
| 1992 | 0.049 | 0.073 | 0.140 | 0.195 | 0.233 | 0.286 | 0.311 | 0.369 | 0.362 | 0.362 |
| 1993 | 0.042 | 0.063 | 0.125 | 0.176 | 0.206 | 0.246 | 0.268 | 0.319 | 0.314 | 0.314 |
| 1994 | 0.036 | 0.054 | 0.113 | 0.160 | 0.186 | 0.219 | 0.241 | 0.292 | 0.285 | 0.285 |
| 1995 | 0.046 | 0.070 | 0.149 | 0.214 | 0.243 | 0.284 | 0.313 | 0.381 | 0.367 | 0.367 |
| 1996 | 0.056 | 0.085 | 0.185 | 0.270 | 0.297 | 0.347 | 0.382 | 0.471 | 0.450 | 0.450 |
| 1997 | 0.055 | 0.084 | 0.188 | 0.279 | 0.300 | 0.349 | 0.382 | 0.473 | 0.452 | 0.452 |
| 1998 | 0.070 | 0.110 | 0.250 | 0.380 | 0.407 | 0.472 | 0.508 | 0.627 | 0.590 | 0.590 |
| 1999 | 0.064 | 0.101 | 0.237 | 0.368 | 0.396 | 0.457 | 0.481 | 0.590 | 0.556 | 0.556 |
| 2000 | 0.074 | 0.118 | 0.279 | 0.444 | 0.496 | 0.575 | 0.587 | 0.703 | 0.664 | 0.664 |
| 2001 | 0.070 | 0.112 | 0.265 | 0.428 | 0.493 | 0.571 | 0.573 | 0.677 | 0.642 | 0.642 |
| 2002 | 0.066 | 0.105 | 0.251 | 0.417 | 0.502 | 0.593 | 0.595 | 0.698 | 0.664 | 0.664 |
| 2003 | 0.068 | 0.108 | 0.262 | 0.438 | 0.541 | 0.632 | 0.625 | 0.704 | 0.666 | 0.666 |


| Year AGe | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{2 0 0 4}$ | 0.070 | 0.111 | 0.271 | 0.461 | 0.590 | 0.689 | 0.686 | 0.750 | 0.709 | 0.709 |
| $\mathbf{2 0 0 5}$ | 0.061 | 0.097 | 0.241 | 0.420 | 0.557 | 0.651 | 0.656 | 0.704 | 0.668 | 0.668 |
| $\mathbf{2 0 0 6}$ | 0.053 | 0.084 | 0.210 | 0.372 | 0.507 | 0.596 | 0.605 | 0.638 | 0.605 | 0.605 |
| $\mathbf{2 0 0 7}$ | 0.050 | 0.080 | 0.198 | 0.356 | 0.502 | 0.602 | 0.625 | 0.657 | 0.626 | 0.626 |
| $\mathbf{2 0 0 8}$ | 0.043 | 0.070 | 0.172 | 0.307 | 0.441 | 0.529 | 0.560 | 0.588 | 0.567 | 0.567 |
| $\mathbf{2 0 0 9}$ | 0.028 | 0.046 | 0.113 | 0.196 | 0.285 | 0.340 | 0.368 | 0.384 | 0.373 | 0.373 |
| $\mathbf{2 0 1 0}$ | 0.020 | 0.034 | 0.081 | 0.137 | 0.199 | 0.236 | 0.258 | 0.264 | 0.258 | 0.258 |
| $\mathbf{2 0 1 1}$ | 0.006 | 0.010 | 0.024 | 0.040 | 0.057 | 0.067 | 0.073 | 0.076 | 0.075 | 0.075 |
| $\mathbf{2 0 1 2}$ | 0.013 | 0.022 | 0.053 | 0.086 | 0.123 | 0.144 | 0.162 | 0.170 | 0.170 | 0.170 |
| $\mathbf{2 0 1 3}$ | 0.022 | 0.038 | 0.094 | 0.152 | 0.217 | 0.252 | 0.285 | 0.305 | 0.304 | 0.304 |
| $\mathbf{2 0 1 4}$ | 0.041 | 0.073 | 0.184 | 0.298 | 0.422 | 0.491 | 0.553 | 0.599 | 0.595 | 0.595 |
| $\mathbf{2 0 1 5}$ | 0.055 | 0.097 | 0.245 | 0.396 | 0.561 | 0.659 | 0.729 | 0.792 | 0.784 | 0.784 |
| $\mathbf{2 0 1 6}$ | 0.049 | 0.088 | 0.220 | 0.357 | 0.508 | 0.609 | 0.670 | 0.731 | 0.723 | 0.723 |
| $\mathbf{2 0 1 7}$ | 0.050 | 0.089 | 0.222 | 0.358 | 0.506 | 0.608 | 0.655 | 0.716 | 0.712 | 0.712 |
| $\mathbf{2 0 1 8}$ | 0.049 | 0.088 | 0.217 | 0.346 | 0.491 | 0.589 | 0.629 | 0.690 | 0.694 | 0.694 |

Table 2.4.4. Blue whiting. Estimated stock numbers-at-age (thousands). Preliminary catch data for 2018 have been used.

| Year <br> Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 3943692 | 3488784 | 4859060 | 2076248 | 2618264 | 2146444 | 1649518 | 1744321 | 1220346 | 2953625 |
| 1982 | 4664381 | 2960440 | 2521907 | 3288160 | 1587728 | 1502456 | 1297429 | 015300 | 891437 | 1940099 |
| 1983 | 18115304 | 3773493 | 1878552 | 1823709 | 1908926 | 1219432 | 1012103 | 853821 | 626640 | 1273072 |
| 1984 | 18014177 | 14418718 | 2439230 | 1234269 | 1263492 | 1392379 | 813203 | 549695 | 481380 | 934332 |
| 1985 | 9611988 | 13503947 | 9735673 | 1453143 | 750716 | 910275 | 745441 | 457726 | 265989 | 724254 |
| 1986 | 7249682 | 6409196 | 9412823 | 5530414 | 943091 | 453109 | 470431 | 376132 | 229802 | 496885 |
| 1987 | 9122396 | 5061002 | 4097616 | 6843814 | 2562277 | 396719 | 253566 | 237646 | 156249 | 292857 |
| 1988 | 6427638 | 6873810 | 3531484 | 2884514 | 3707908 | 1260555 | 198998 | 125516 | 99236 | 171477 |
| 1989 | 8537748 | 4633873 | 4991355 | 2430816 | 2129918 | 1682807 | 352575 | 102615 | 60199 | 115895 |
| 1990 | 18736561 | 6005399 | 3106085 | 2737879 | 1482622 | 1189410 | 563082 | 121621 | 33337 | 84487 |
| 1991 | 8988548 | 15608671 | 4282592 | 1798497 | 1492165 | 870272 | 561713 | 190059 | 32994 | 45240 |
| 1992 | 6713109 | 7409881 | 12477451 | 3311350 | 1265977 | 794944 | 487709 | 288632 | 102008 | 39508 |
| 1993 | 4998700 | 5132861 | 5288210 | 9706302 | 2261567 | 978157 | 518320 | 283539 | 157823 | 74937 |
| 1994 | 8135997 | 3417183 | 4076316 | 3410694 | 6919250 | 1441441 | 764562 | 328596 | 205616 | 117856 |
| 1995 | 9335808 | 5886126 | 3141338 | 2577034 | 2857853 | 3751871 | 1039398 | 543237 | 219866 | 185984 |
| 1996 | 27984503 | 7110350 | 4084034 | 2399449 | 1558932 | 1867109 | 2242241 | 644926 | 306700 | 248522 |
| 1997 | 44654015 | 2127648 | 5494343 | 2574210 | 1424080 | 1071777 | 1064883 | 1218157 | 290420 | 333352 |
| 1998 | 26698034 | 37684012 | 16390067 | 3500268 | 1380547 | 928845 | 780824 | 604511 | 618616 | 294523 |
| 1999 | 20324984 | 20540726 | 27575781 | 10528634 | 1715494 | 776468 | 521586 | 410706 | 238132 | 428401 |
| 2000 | 39079520 | 15297659 | 16594148 | 15830032 | 4347296 | 1108867 | 472166 | 324122 | 155373 | 314533 |
| 2001 | 55497740 | 31482540 | 12072473 | 10747410 | 7479057 | 1704442 | 491788 | 227543 | 161849 | 180036 |
| 2002 | 48380225 | 45012302 | 20420467 | 8317975 | 5471791 | 3413372 | 695587 | 256524 | 102664 | 154653 |
| 2003 | 52143422 | 38633073 | 34816368 | 13565168 | 5063018 | 2969554 | 1214175 | 348189 | 90256 | 107123 |
| 2004 | 27840934 | 41597583 | 29708160 | 20801295 | 7277885 | 2454603 | 1318032 | 508010 | 152905 | 81304 |
| 2005 | 21854787 | 20958087 | 28217872 | 17943922 | 10759639 | 3244168 | 1107069 | 515249 | 193741 | 99173 |
| 2006 | 8915416 | 15214984 | 21736797 | 19133061 | 9422615 | 4457765 | 1359644 | 483782 | 218294 | 120636 |
| 2007 | 4873765 | 5916804 | 12980388 | 15742899 | 10276454 | 4686246 | 1836896 | 611595 | 228534 | 162864 |
| 2008 | 5730029 | 3456510 | 4319952 | 10977175 | 9140533 | 4906009 | 1864245 | 762892 | 237690 | 198625 |
| 2009 | 5577076 | 3930396 | 2416993 | 3700656 | 6945693 | 4727082 | 2206941 | 858654 | 327481 | 190213 |
| 2010 | 14802057 | 4904170 | 2336334 | 1860322 | 3357171 | 4327255 | 2843716 | 1206054 | 417629 | 268918 |
| 2011 | 18517015 | 12923098 | 3271120 | 1650914 | 1613891 | 2585190 | 2684504 | 1362763 | 816029 | 394679 |
| 2012 | 18200627 | 14807332 | 12169939 | 2276802 | 1178572 | 1603821 | 2319215 | 2089712 | 1076181 | 899308 |
| 2013 | 15173469 | 15156414 | 11246027 | 7226372 | 2191295 | 1069671 | 1355113 | 1601432 | 1329327 | 1363701 |
| 2014 | 34951221 | 12049711 | 13315826 | 7775563 | 4282644 | 1315215 | 905664 | 962744 | 984953 | 1461751 |


| Year AGE | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 57777072 | 30998427 | 10369579 | 8225393 | 4084705 | 1678424 | 708618 | 491099 | 454668 | 1011553 |
| 2016 | 29276819 | 51763155 | 20193887 | 7359970 | 4105291 | 1695049 | 657303 | 320562 | 199265 | 543064 |
| 2017 | 9104584 | 23379038 | 41023710 | 14373309 | 4332574 | 1911764 | 656796 | 240411 | 135352 | 321343 |
| 2018 | 11037772 | 6697833 | 18011646 | 26425229 | 8372176 | 2411845 | 738959 | 236217 | 102601 | 209397 |
| 2019 |  | 8604992 | 5024187 | 11875347 | 15302312 | 4195934 | 1096200 | 322551 | 97019 | 127623 |

Table 2.4.5. Blue whiting. Estimated recruitment in thousands, spawning-stock biomass (SSB) in tonnes, average fishing mortality for ages 3 to 7 (F3-7) and total-stock biomass (TBS) in tonnes. Preliminary catch data for 2018 are included.

Year R(age 1) Low High ssb Low high | Fbar |
| :---: |
| $(3-7)$ | Low high tsb Low high

| 1981 | 3943692 | 2520198 | 6171224 | 2843780 | 2221416 | 3640508 | 0.258 | 0.186 | 0.359 | 3341965 | 2661374 | 4196604 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 4664381 | 2946828 | 7383007 | 2303718 | 1821256 | 2913988 | 0.220 | 0.162 | 0.301 | 2772882 | 2232015 | 3444814 |
| 1983 | 18115304 | 11695267 | 28059578 | 1858444 | 1502264 | 2299072 | 0.255 | 0.190 | 0.342 | 2882436 | 2334841 | 3558459 |
| 1984 | 18014177 | 11740672 | 27639865 | 1752199 | 1441049 | 2130531 | 0.320 | 0.241 | 0.423 | 3080871 | 2473661 | 3837132 |
| 1985 | 9611988 | 6291195 | 14685653 | 2088626 | 1714249 | 2544764 | 0.357 | 0.273 | 0.468 | 3226731 | 2620479 | 3973239 |
| 1986 | 7249682 | 4776823 | 11002687 | 2271261 | 1867669 | 2762068 | 0.437 | 0.335 | 0.569 | 3113127 | 2566429 | 3776281 |
| 1987 | 9122396 | 5997119 | 13876349 | 1931350 | 1590580 | 2345128 | 0.420 | 0.321 | 0.549 | 2817310 | 2325903 | 3412539 |
| 1988 | 6427638 | 4223067 | 9783063 | 1637240 | 1359781 | 1971313 | 0.440 | 0.337 | 0.575 | 2426489 | 2011389 | 2927256 |
| 1989 | 8537748 | 5588110 | 13044327 | 1547528 | 1289249 | 1857550 | 0.528 | 0.406 | 0.687 | 2395234 | 1976026 | 2903376 |
| 1990 | 18736561 | 12078959 | 29063658 | 1360082 | 1122934 | 1647312 | 0.517 | 0.390 | 0.684 | 2500723 | 1988528 | 3144845 |
| 1991 | 8988548 | 5728713 | 14103339 | 177945 | 1420075 | 222979 | 0.290 | 0.211 | 0.397 | 3221695 | 2508294 | 4138000 |
| 1992 | 6713109 | 4333876 | 10398504 | 2459581 | 1936174 | 3124481 | 0.233 | 0.170 | 0.320 | 3529552 | 2782205 | 4477648 |
| 1993 | 4998700 | 3187939 | 7837980 | 2540869 | 2009014 | 321352 | 0.204 | 0.149 | 0.279 | 3420516 | 2724355 | 4294568 |
| 1994 | 8135997 | 5237893 | 12637610 | 253536 | 2026669 | 3171744 | 0.184 | 0.134 | 0.252 | 3418580 | 2759310 | 4235365 |
| 1995 | 9335808 | 6074990 | 14346906 | 2312732 | 1891556 | 2827686 | 0.240 | 0.179 | 0.322 | 3361400 | 2751602 | 4106337 |
| 1996 | 27984503 | 18249895 | 42911611 | 2212106 | 1827081 | 2678269 | 0.296 | 0.223 | 0.394 | 3728177 | 3018014 | 4605447 |
| 1997 | 44654015 | 29188410 | 68314138 | 2467411 | 2034689 | 2992161 | 0.299 | 0.226 | 0.397 | 5427534 | 4244969 | 6939538 |
| 1998 | 26698034 | 17560241 | 40590846 | 3674788 | 2986783 | 4521274 | 0.403 | 0.308 | 0.528 | 6810285 | 5413990 | 8566692 |
| 1999 | 20324984 | 13302894 | 31053766 | 4438043 | 3591370 | 5484320 | 0.388 | 0.295 | 0.509 | 7166282 | 5792673 | 8865611 |
| 2000 | 39079520 | 25520182 | 59843181 | 4235751 | 3497561 | 5129742 | 0.476 | 0.367 | 0.619 | 7456428 | 6049105 | 9191166 |
| 2001 | 55497740 | 36531341 | 84311145 | 4571904 | 3792756 | 5511112 | 0.466 | 0.358 | 0.606 | 8985204 | 7212545 | 11193538 |
| 2002 | 48380225 | 31821100 | 73556421 | 5397878 | 4469870 | 6518552 | 0.472 | 0.362 | 0.615 | 10294085 | 8290040 | 12782590 |
| 2003 | 52143422 | 34804230 | 78120861 | 6834267 | 5636111 | 8287133 | 0.500 | 0.388 | 0.643 | 11752059 | 9580906 | 14415221 |
| 2004 | 27840934 | 18409030 | 42105293 | 6730016 | 5610974 | 8072237 | 0.539 | 0.422 | 0.690 | 10293010 | 8540046 | 12405795 |
| 2005 | 21854787 | 14554141 | 32817581 | 5976941 | 4982409 | 7169991 | 0.505 | 0.392 | 0.650 | 8402061 | 6986988 | 10103727 |
| 2006 | 8915416 | 5871058 | 13538385 | 5824060 | 4830917 | 7021375 | 0.458 | 0.353 | 0.594 | 7639999 | 6342177 | 9203399 |
| 2007 | 4873765 | 3198282 | 7426981 | 4641908 | 3837586 | 5614809 | 0.456 | 0.348 | 0.599 | 5668818 | 4697569 | 6840878 |
| 2008 | 5730029 | 3706275 | 8858823 | 3584601 | 2920250 | 4400090 | 0.402 | 0.298 | 0.543 | 4396217 | 3597370 | 5372459 |
| 2009 | 5577076 | 3482285 | 8932002 | 2755802 | 2185478 | 3474958 | 0.260 | 0.187 | 0.362 | 3459138 | 2762543 | 4331386 |
| 2010 | 14802057 | 9497910 | 23068329 | 2676765 | 2079551 | 3445491 | 0.182 | 0.128 | 0.260 | 3720075 | 2917079 | 4744115 |
| 2011 | 18517015 | 11981809 | 28616700 | 2681413 | 2093175 | 3434960 | 0.052 | 0.035 | 0.078 | 4352159 | 3402862 | 5566282 |
| 2012 | 18200627 | 11951951 | 27716214 | 3379349 | 2706766 | 4219058 | 0.113 | 0.083 | 0.154 | 4982053 | 3987711 | 6224336 |
| 2013 | 15173469 | 9976381 | 23077923 | 3665883 | 2991885 | 4491716 | 0.200 | 0.150 | 0.267 | 5396783 | 4383191 | 6644762 |
| 2014 | 34951221 | 22594064 | 54066761 | 3870679 | 3185710 | 4702926 | 0.389 | 0.295 | 0.514 | 6361361 | 5125400 | 7895367 |
| 2015 | 57777072 | 36854356 | 90577896 | 4001425 | 3254742 | 4919409 | 0.518 | 0.395 | 0.680 | 7695209 | 5986917 | 9890940 |
| 2016 | 29276819 | 17996990 | 47626416 | 4575498 | 3546395 | 5903228 | 0.473 | 0.345 | 0.649 | 8320101 | 6256088 | 11065074 |
| 2017 | 9104584 | 5101379 | 16249224 | 5508728 | 4013548 | 7560914 | 0.470 | 0.315 | 0.702 | 7815325 | 5645224 | 10819641 |


| Year R(AGe 1) | Low | High | SSB | Low | High | Fbar <br> $(3-7)$ | Low High | TSB | Low High |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2018 | 11037772 | 4815497 | 25300070 | 5422226 | 3586596 | 8197337 | 0.454 | 0.262 | 0.787 | 6952013 | 4582975 |

*assuming long tem GM(1981-2017) recruitment (14580847)

Table 2.4.6 .Blue whiting. Model estimate of total catch weight (in tonnes) and Sum of Product of catch number and mean weight at age for ages 1-10+ (Observed catch). Preliminary catch data for 2018 are included.

| Year | Estimate | Low | High | Observed catch |
| :---: | :---: | :---: | :---: | :---: |
| 1981 | 784934 | 556689 | 1106761 | 922980 |
| 1982 | 543231 | 409101 | 721340 | 550643 |
| 1983 | 512673 | 392727 | 669252 | 553344 |
| 1984 | 562099 | 430053 | 734690 | 615569 |
| 1985 | 638465 | 497039 | 820132 | 678214 |
| 1986 | 760762 | 592701 | 976476 | 847145 |
| 1987 | 638856 | 497827 | 819838 | 654718 |
| 1988 | 568654 | 443878 | 728504 | 552264 |
| 1989 | 618626 | 486149 | 787205 | 630316 |
| 1990 | 553978 | 432532 | 709524 | 558128 |
| 1991 | 406313 | 313193 | 527120 | 364008 |
| 1992 | 438851 | 342851 | 561731 | 474592 |
| 1993 | 439997 | 342045 | 565999 | 475198 |
| 1994 | 424834 | 328214 | 549897 | 457696 |
| 1995 | 507850 | 399237 | 646013 | 505176 |
| 1996 | 597801 | 470210 | 760014 | 621104 |
| 1997 | 641181 | 500273 | 821778 | 639681 |
| 1998 | 1077597 | 835390 | 1390027 | 1131955 |
| 1999 | 1243881 | 958805 | 1613717 | 1261033 |
| 2000 | 1503366 | 1167941 | 1935124 | 1412449 |
| 2001 | 1562271 | 1214242 | 2010053 | 1771805 |
| 2002 | 1716431 | 1333431 | 2209440 | 1556955 |
| 2003 | 2194967 | 1713850 | 2811144 | 2365319 |
| 2004 | 2314871 | 1814965 | 2952469 | 2400795 |
| 2005 | 1996040 | 1567578 | 2541613 | 2018344 |
| 2006 | 1841398 | 1444963 | 2346596 | 1956239 |
| 2007 | 1545598 | 1210851 | 1972889 | 1612269 |
| 2008 | 1162864 | 904241 | 1495455 | 1251851 |


| Year | Estimate | Low | High | ObSERVED CATCH |
| :--- | :---: | :---: | :---: | :---: |
| 2009 | 654781 | 508179 | 843676 | 634978 |
| 2010 | 477149 | 364441 | 624713 | 539539 |
| 2011 | 136638 | 99573 | 187499 | 103771 |
| 2012 | 329201 | 258410 | 419384 | 375692 |
| 2013 | 592951 | 464578 | 756796 | 613863 |
| 2014 | 1108159 | 861440 | 1425539 | 1147650 |
| 2015 | 1355520 | 1064908 | 1725440 | 1390656 |
| 2016 | 1267922 | 991376 | 1621611 | 1180786 |
| 2017 | 1510357 | 1180296 | 1932717 | 1555069 |
| 2018 | 1688894 | 1292561 | 2206752 | 1712874 |

Table 2.8.1.1.Blue whiting. Input to short-term projection (median values for exploitation pattern and stock numbers).

| AGE | MEAN <br> WEIGHT IN <br> THE STOCK <br> (KG) | MEAN WEIGHT IN <br> THE CATCH (KG) | PROPORTION <br> MATURE | NATURAL <br> MORTALITY | EXPLOI- <br> TATION <br> PATTERN | STOCK <br> NUMBER(2019) <br> (THOUSANDS) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age 1 | 0.047 | 0.047 | 0.11 | 0.20 | 0.108 | 14580847 |
| Age 2 | 0.069 | 0.069 | 0.40 | 0.20 | 0.193 | 8604992 |
| Age 3 | 0.086 | 0.086 | 0.82 | 0.20 | 0.477 | 5024187 |
| Age 4 | 0.106 | 0.106 | 0.86 | 0.20 | 0.762 | 11875347 |
| Age 5 | 0.125 | 0.125 | 0.91 | 0.20 | 1.080 | 15302312 |
| Age 6 | 0.143 | 0.143 | 0.94 | 0.20 | 1.296 | 4195934 |
| Age 7 | 0.158 | 0.158 | 1.00 | 0.20 | 1.385 | 1096200 |
| Age 8 | 0.174 | 0.174 | 1.00 | 0.20 | 1.519 | 322551 |
| Age 9 | 0.189 | 0.189 | 1.00 | 0.20 | 1.528 | 97019 |
| Age 10 | 0.206 | 0.206 | 1.00 | 0.20 | 1.528 | 127623 |

Table 2.8.2.1.1. Blue whiting. Deterministic forecast, intermediate year assumptions and recruitments.

| Values | Value | Notes |
| :--- | :--- | :--- |
| F ages 3-7 (2018) | 0.454 | From the assessment (preliminary 2018 catches) |
| SSB (2019) | 4326857 | From forecast |
| R age 1 (2018) | 11037772 | From assessment |
| R age 1 (2019) | 14580847 | GM (1981-2017) |
| R age $1(2020)$ | 14580847 | GM (1981-2017) |
| Total catch (2018) | 1712874 | Preliminary 2018 catches as estimated by the WG, <br> based on declaredquotas and expecteduptake. |

Table 2.8.2.2.1.Blue whiting. Deterministic forecast(weights in tonnes).

| BASIS | CATCH (2019) | $F(2019)$ | $\begin{gathered} \text { SSB } \\ (2020) \end{gathered}$ | $\begin{gathered} \text { \% SSB } \\ \text { CHANGE* } \end{gathered}$ | \% CATCH <br> CHANGE** | \% Advice <br> CHANGE*** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY approach: FMSY | 1143629 | 0.320 | 3752236 | -13.3 | -33.2 | -17.6 |
| $\mathrm{F}=0$ | 4 | 0.000 | 4850444 | 12.1 | -100.0 | -100.0 |
| Fpa | 1725357 | 0.530 | 3201021 | -26.0 | 0.7 | 24.3 |
| Flim | 2476742 | 0.880 | 2499796 | -42.2 | 44.6 | 78.5 |
| SSB (2020) = Blim | 3587714 | 1.735 | 1500171 | -65.3 | 109.5 | 158.5 |
| SSB (2020 = Bpa | 2747920 | 1.039 | 2250714 | -48.0 | 60.4 | 98.0 |
| SSB (2020) = MSY Btrigger | 2747920 | 1.039 | 2250714 | -48.0 | 60.4 | 98.0 |
| $\mathrm{F}=\mathrm{F}$ (2018) | 1528542 | 0.454 | 3386825 | -21.7 | -10.8 | 10.1 |
| SSB (2020) = SSB (2019) | 544778 | 0.140 | 4325259 | -0.0 | -68.2 | -60.7 |
| Catch (2019) = Catch (2018) | 1712790 | 0.525 | 3212862 | -25.7 | -0.0 | 23.4 |
| Catch (2019) = Catch (2018) -20\% | 1370342 | 0.397 | 3536701 | -18.3 | -20.0 | -1.3 |
| Catch (2019) = Advice (2018) -20 \% | 1109872 | 0.309 | 3784400 | -12.5 | -35.2 | -20.0 |
| $\mathrm{F}=0.05$ | 202887 | 0.050 | 4654469 | 7.6 | -88.2 | -85.4 |
| $\mathrm{F}=0.1$ | 396102 | 0.100 | 4468252 | 3.3 | -76.9 | -71.5 |
| $\mathrm{F}=0.15$ | 580153 | 0.150 | 4291275 | -0.8 | -66.1 | -58.2 |
| $\mathrm{F}=0.16$ | 615906 | 0.160 | 4256945 | -1.6 | -64.0 | -55.6 |
| $\mathrm{F}=0.17$ | 651316 | 0.170 | 4222960 | -2.4 | -62.0 | -53.1 |
| $\mathrm{F}=0.18$ | 686385 | 0.180 | 4189318 | -3.2 | -59.9 | -50.5 |
| $\mathrm{F}=0.19$ | 721119 | 0.190 | 4156014 | -3.9 | -57.9 | -48.0 |
| $\mathrm{F}=0.2$ | 755520 | 0.200 | 4123044 | -4.7 | -55.9 | -45.6 |
| $\mathrm{F}=0.21$ | 789591 | 0.210 | 4090406 | -5.5 | -53.9 | -43.1 |
| $\mathrm{F}=0.22$ | 823337 | 0.220 | 4058095 | -6.2 | -51.9 | -40.7 |
| $\mathrm{F}=0.23$ | 856761 | 0.230 | 4026108 | -7.0 | -50.0 | -38.3 |
| $\mathrm{F}=0.24$ | 889866 | 0.240 | 3994442 | -7.7 | -48.0 | -35.9 |
| $\mathrm{F}=0.25$ | 922655 | 0.250 | 3963093 | -8.4 | -46.1 | -33.5 |
| $\mathrm{F}=0.26$ | 955133 | 0.260 | 3932057 | -9.1 | -44.2 | -31.2 |
| $\mathrm{F}=0.27$ | 987302 | 0.270 | 3901331 | -9.8 | -42.4 | -28.9 |
| $\mathrm{F}=0.28$ | 1019165 | 0.280 | 3870912 | -10.5 | -40.5 | -26.6 |
| $\mathrm{F}=0.29$ | 1050726 | 0.290 | 3840796 | -11.2 | -38.7 | -24.3 |
| $\mathrm{F}=0.3$ | 1081989 | 0.300 | 3810980 | -11.9 | -36.8 | -22.0 |
| $\mathrm{F}=0.31$ | 1112955 | 0.310 | 3781462 | -12.6 | -35.0 | -19.8 |
| $\mathrm{F}=0.32$ | 1143629 | 0.320 | 3752236 | -13.3 | -33.2 | -17.6 |
| $\mathrm{F}=0.33$ | 1174013 | 0.330 | 3723302 | -13.9 | -31.5 | -15.4 |
| $\mathrm{F}=0.34$ | 1204111 | 0.340 | 3694655 | -14.6 | -29.7 | -13.2 |
| $\mathrm{F}=0.35$ | 1233925 | 0.350 | 3666292 | -15.3 | -28.0 | -11.1 |


| Year | Estimate | Low | High | ObSERVED CATCh |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: |
| $\mathrm{F}=0.45$ |  | 1517116 | 0.450 | 3397635 | -21.5 | -11.4 | 9.3 |
| $\mathrm{~F}=0.5$ |  | 1649075 | 0.500 | 3272945 | -24.4 | -3.7 | 18.8 |

Weights in tonnes.
*) SSB 2020 relative to SSB 2019.
**) Catch 2019 relative to expected catch in 2018 ( 1712874 tonnes).
***) Catch 2019 relative to advice for 2018 (1387872 tonnes).

### 2.16 Figures

## WHB catch 2017

1540077 tonnes in total 200 m and 1000 m depth contours in blue


Figure 2.2.1. Blue whiting landings (ICES estimates) in 2017 by ICES rectangle. The 200 m and 1000 $m$ depth contours are indicated in blue. The catches on the map constitute $98.8 \%$ of the total landings.

WHB catch 2017 First quarter 653423 tonnes, 42.4\%


Third quarter 64911 tonnes, $\mathbf{4 . 2 \%}$

200m and 1000 m depth contours in blue Second quarter 693376 tonnes, 45\%


Fourth quarter 128367 tonnes, 8.3\%



Figure 2.2.2. Blue whiting total catches pr quarter (ICES estimates) 2017 by ICES rectangle. The 200 m and 1000 m depth contours are indicated in blue. The catches on the map constitute $98.8 \%$ of the total landings.


Figure 2.3.1.1. Blue whiting. ICES estimated catches (tonnes) in 2017 by ICES division area and country.

A


B


Figure 2.3.1.2. Blue whiting. (A) ICES estimated catches (tonnes) of blue whiting by fishery subareas from 1988-2017 and (B) the percentage contribution to the overall catch by fishery subarea over the same period.


Figure 2.3.1.3. Blue whiting. Distribution of 2017 ICES estimated catches (in percentage) by ICES division area.


Figure 2.3.1.4. Blue whiting. Distribution of 2017 ICES estimated catches (in percentage) by quarter.


Figure 2.3.1.5. Blue whiting. Distribution of 2017 ICES estimated catches (tonnes) by country and by quarter.


Figure 2.3.1.6. Blue whiting. Distribution of 2017 ICES estimated catches (tonnes) by ICES division area and by quarter.


Figure 2.3.1.7. Blue whiting. Catch-at-age numbers (CANUM) distribution by quarter and ICES division area for 2017.


Figure 2.3.1.8. Blue whiting. 2017 ICES catches (tonnes) sampled and estimated by ICES division area.

2.3.1.2.1 .Blue whiting. Length ( cm ) for 2017 ICES estimated catches (tonnes).This length distribution represents only $67 \%$ of the 2017 ICES estimated total catches (tonnes).


Figure 2.3.1.3.1. Blue whiting. Mean length (mm) by age ( $0-10$ year), by quarter ( $1,2,4$ ), by country for ICES division area 27.6.a. These data only comprises the 2017 ICES catch-at-age sampled estimates for ICES division area 27.6.a.


Figure 2.3.2.1. Blue whiting. Distribution of 2018 preliminary catches (tonnes) by ICES division area and quarter.


Figure 2.3.3.1. Blue whiting. Catch proportion at age, 1981-2018. Preliminary values for 2018 have been used.


Figure 2.3.3.2. Blue whiting. Age disaggregated catch (numbers) plotted on $\log$ scale. The labels for each panel indicate year classes. The grey dotted lines correspond to $Z=0.6$. Preliminary catch-atage for 2018 have been used.


Figure 2.3.4.1. Blue whiting. Mean catch (and stock) weight (kg) at age by year. Preliminary values for 2018 (average of 2015-2017) have been used.


Figure 2.3.7.1.1. Blue whiting. (A) Estimate of total biomass from the International blue whiting spawning stock survey. The black dots and error bands are StoX estimates with $95 \%$ confidence intervals. (B) Internal consistency within the International blue whiting spawning stock survey. The upper left part of the plots shows the relationship between log index-at-age within a cohort. Linear regression line shows the best fit to the log-transformed indices. The lower-right part of the plots shows the correlation coefficient (r) for the two ages plotted in that panel. The background colour of each panel is determined by the $r$ value, where red equates to $r=1$ and white to $r<0$.


2015
2016


Figure 2.3.7.1.2. Map of blue whiting acoustic density ( $\mathrm{sA}, \mathrm{m} 2 / \mathrm{nm} 2$ ) found during the spawning survey in spring 2015-2018.


Figure 2.3.7.1.3. Blue whiting. Length (line) and age (bars) distribution of the blue whiting stock in the area to the west of the British Isles, spring 2014 (lower panel) to 2018 (upper panel). Spawningstock biomass and numbers are given.


Figure 2.3.7.2.. 1 Blue whiting spatial distribution according to NASC values allocated to this species during PELACUS-IBWSS 0318.


Figure 2.3.7.2.2. Blue whiting length distribution as estimated during PELACUS-IBWSS 0318.


Figure 2.4.1. Blue Whiting. OSA (One Step Ahead) residuals (see Berg and Nielsen, 2016) from catch-at-age and the IBWSS survey. Red (lighter) bubbles show that the observed value is less than the expected value. Preliminary catch data for 2018 have been used.


Figure 2.4.2. Blue whiting. Joint sample residuals (Process errors) for stock number and F at age. Red (lighter) bubbles show that the observed value is less than the expected value. Preliminary catch data for 2018 have been used.

## Residual catch



IBWSS


Figure 2.4.3. Blue whiting. The correlation matrix between ages for the catches and survey indices. Each ellipse represents the level curve of a bivariate normal distribution with the corresponding correlation. Hence, the sign of a correlation corresponds to the sign of the slope of the major ellipse axis. Increasingly darker shading is used for increasingly larger absolute correlations, while uncorrelated pairs of ages are depicted as circles with no shading.


Figure 2.4.4. Blue whiting. $F$ at age and exploitation pattern ( $F$ scaled to mean $F$ all ages, and $F$ scaled to mean $F$ ages 3-7). Values for 2018 are preliminary.


Figure 2.4.5. Blue whiting. Retrospective analysis of recruitment (age 1), SSB (tonnes), F and total catch using the SAM model. The $95 \%$ confidence interval is shown for the most recent assessment.


Figure 2.4.6.Blue whiting. SAM final run: Stock summary, total catches (tonnes), recruitment (age 1), F and SSB (tonnes). The graphs show the median value and the $95 \%$ confidence interval. The catch plot does also include the observed catches (x). Catches for 2018 are preliminary.


Figure 2.4.1.1. Blue whiting. Comparison of SSB, F and recruitment estimated by the assessment programs XSA, TISVPA and SAM. Catch values for 2018 are preliminary.


Figure 2.8.1.1. Blue whiting young fish indices from five different surveys and recruitment index from the assessment, standardized by dividing each series by their mean. BarSea - Norwegian bot-tom-trawl survey in the Barents Sea, IESNS: International Ecosystem Survey in the Nordic Seas in May ( 1 and 2 is the age groups), IBWSS: International Blue Whiting Spawning Stock survey (1 and 2 is the age groups), FO: the Faroese bottom-trawl surveys in spring, IS: the Icelandic bottom-trawl survey in spring, SAM: recruits from the assessment.


Figure 2.9.1. Blue whiting. Comparison of the 2010-2018 assessments.

## 3 Northeast Atlantic boarfish (Capros aper)

The boarfish (Capros aper, Linnaeus) is a deep bodied, laterally compressed, pelagic shoaling species distributed from Norway to Senegal, including the Mediterranean, Azores, Canaries, Madeira and Great Meteor Seamount (Blanchard \& Vandermeirsch 2005).

Boarfish is targeted in a pelagic trawl fishery for fish meal, to the southwest of Ireland. The boarfish fishery is conducted primarily in shelf waters and the first landings were reported in 2001. Landings were at very low levels from 2001-2005. The main expansion period of the fishery was 2006-2010 when unrestricted landings increased from 2 772 t to 137503 t . A restrictive TAC of 33000 t was implemented in 2011. In 2011, ICES was asked by the European Commission to provide advice for 2012. In 2018, ICES has been considering this stock for 8 years.

An analysis of bottom trawl survey data suggests a continuity of distribution spanning ICES Subareas 27.4, 6, 7, 8 and 9 (Figure 3.1). Isolated occurrences appear in the North Sea (ICES Subarea 27.4) in some years indicating spill-over into this region. A hiatus in distribution was suggested between ICES Divisions 27.8.c and 9.a as boarfish were considered very rare in northern Portuguese waters but abundant further south (Cardador \& Chaves 2010). Results from a dedicated genetic study on the stock structure of boarfish within the Northeast Atlantic and Mediterranean Sea suggests that this hiatus represents a true stock separation (Farrell et al. (2016); see section 3.11). Based on these data, a single stock is considered to exist in ICES Subareas 27.4, 6, 7, 8 and the northern part of 9.a. This distribution is slightly broader than the current EC TAC area (27.6, 7 and 8) and for the purposes of assessment in 2018 only data from these areas were utilized.

### 3.1 The fishery

### 3.1.1 Advice and management applicable from 2011 to 2018

In 2011 a TAC was set for this species for the first time, covering ICES Subareas 6, 7 and 8. This TAC was set at 33000 t . Before 2010, the fishery was unregulated. In October 2010, the European Commission notified national authorities that under the terms of Annex 1 of Regulation 850/1998, industrial fisheries for this species should not proceed with mesh sizes of less than 100 mm . In 2011, the European Parliament voted to change Regulation 850/1998 to allow fishing using mesh sizes ranging from 32 to 54 mm .

For 2012, ICES advised that catches of boarfish should not increase, based on precautionary considerations. As supporting information, ICES noted that it would be cautious that landings did not increase above 82000 t , the average over the period 20082010, during which the stock did not appear to be overexploited. In 2012 the TAC was set at 82000 t by the Council of the European Union.

For 2013, ICES advised that catches of boarfish should not be more than 82000 t . This was based on applying a harvest ratio of $12.2 \%$ (F0.1, as an FmSY proxy). For 2013, the TAC was set at 82000 t by the Council of the European Union.

For 2014, ICES advised that, based on FMSY (0.23), catches of boarfish should not be more than 133957 t , or 127509 t when the average discard rate of the previous ten years ( 6448 t ) is taken into account. For 2014 the TAC was set at 133957 t by the Council of
the European Union. This advice was based on a Schaefer state space surplus production model (see section 3.6.3 for further details).

In 2014 there was concern about the use of the production model (see stock annex). ICES considered that the model was no longer suitable for providing category 1 advice and further model development was required. The model is still considered suitable for category 3 advice. The advised catch for 2015 of 53296 t was based on the data limited stock HCR and an index calculated (method 3.1; ICES, 2012) using the total stock biomass trends from the model. Further work has been undertaken in 2015 to address the issues with the surplus production model and this work has been continued since then.

For 2016, ICES advised based on the precautionary approach that catches should be no more than 42637 t .

For 2017, ICES advised based on the precautionary approach that catches should be no more than 27288 t . For the first time, the precautionary buffer has been applied resulting in a $36 \%$ reduction compared to the year before. The acoustic survey suggested that the stock abundance was at an historic low.

In 2017, the Advice Drafting Group decided the advice of 21830 proposed ( $20 \%$ reduction) would stand for 2 years. The assessment run in 2018 confirms that the biomass is rather stable and at a low level.

Since 2011, there has been a provision for bycatch of boarfish (also whiting, haddock and mackerel) to be taken from the Western and North Sea horse mackerel EC quotas. These provisions are shown in the text table below. The effect of this is that a quantity not exceeding the value indicated of these 4 species combined may be landed legally and subtracted from quotas for horse mackerel.

| Year | North.SEA.(T) | Western.(T) |
| :---: | :---: | :---: |
| 2011 | 2031 | 7779 |
| 2012 | 2148 | 7829 |
| 2013 | 1702 | 7799 |
| 2014 | 1392 | 5736 |
| 2015 | 583 | 4202 |
| 2016 | 760 | 5443 |
| 2017 | 912 | 4191 |
| 2018 | 759 | 5053 |

In 2010, an interim management plan was proposed by Ireland, which included a number of measures to mitigate potential bycatch of other TAC species in the boarfish fishery. A closed season from the 15th March to 31st August was proposed, as anecdotal evidence suggests that mackerel and boarfish are caught in mixed aggregations during this period. A closed season was proposed in ICES Division 7.g from 1st September to 31st October, in order to prevent catches of Celtic Sea herring, which is known to form feeding aggregations in this region at these times. Finally, if catches of a species covered by a TAC, other than boarfish, amount to more than $5 \%$ of the total catch by day by ICES statistical rectangle, then fishing must cease in that rectangle for 5 days.

In August 2012 the Pelagic RAC proposed a long term management plan for boarfish (see section 3.15). The management plan was not fully evaluated by ICES. However, in 2013, ICES advised that Tier 1 of the plan can be considered precautionary if a Category 1 assessment is available.

A revised draft management strategy was proposed by the Pelagic AC in July 2015. This management strategy aims to achieve exploitation of boarfish in line with the precautionary approach to fisheries management, FAO guidelines for new and developing fisheries, and the ICES form of advice. ICES evaluated the plan and considered it to be precautionary, in that that it follows the rationale for TAC setting enshrined in the ICES advice, but with additional caution.

The closed season, in the interim and revised management plans, has been enacted in legislation in Ireland, though not other countries.

### 3.1.2 The fishery in recent years

The first landings of boarfish were reported in 2001. Landings fluctuated between 100 and 700 t per year up to 2005 (Tables 3.1.2.1 \& 3.1.2.2). In 2006 the landings began to increase considerably as a target fishery developed. Cumulative landings since 2001 are now over 500000 t . The fishery targets dense shoals of boarfish from September to March. Catches are generally free from bycatch from September to February. From March onward a bycatch of mackerel can be found in the catches and the fishery generally ceases at this time. Information on the bycatch of other species in the boarfish fishery is sparse, though thought to be minimal. The fishery uses typical pelagic pair trawl nets with mesh sizes ranging from 32 to 54 mm . Preliminary information suggests that only the smallest boarfish escape this gear.

From 2001 to 2006 only Ireland reported landings of boarfish. In 2007 UK-Scotland reported landings of 772 t . Scottish landings peaked at 9241 t in 2010 and have declined since then with no fishery in 2015. Denmark joined the fishery in 2008 and landed 3098 t . Danish landings increased to 39805 t in 2010 but have declined considerably to only 29 t in 2015 and were null in 2016 and 2017. The vast majority of catches have come from ICES Division 27.7.j and 27.7.h (Figure 3.1.2.1 and Table 3.1.2.1). Since 2011 landings have been regulated by a TAC.

In 2014 and subsequent years, the TAC has not been caught. This is thought to be partly due to lesser availability of fishable aggregations, and partly due to economic and administrative reasons. According to the industry, fishable aggregations were not always available during the fishery. The season coincides with the mackerel and horse mackerel fisheries. Also, the Irish quota was allocated to individual boats, with non-specialist vessels receiving allocations that were not used.

In 2015 Q3 and Q4 individual boat quotas have been removed in Ireland, in an attempt to allow the specialist 6-7 vessels to target the stock without (what the industry considers to be unnecessary) constraints. The same year, the Netherlands ( 375 t ), UK England (104 t) and Germany ( 4 t ) reported boarfish landings for the first time. These landings were mainly bycatch from freezer trawlers.

In 2016 a total of 19315 t of boarfish were caught (Table 3.1.2.1). Ireland continued to be the main participant taking 17496 t but is below its 29464 quota. Denmark took only 337 t , significantly under its national quota of 10463 t . Scotland reported no boarfish landings. Table 3.1.2.2 shows that two thirds of the Irish landings were taken in ICES divisions 7.h and 8.a. Thirty-two Irish registered fishing vessels reported catches with the majority made in Q1 (7143t) and Q4 (8711t).

Previous to the development of the target fishery, boarfish was a discarded bycatch in pelagic fisheries for mackerel in ICES Subareas 7 and 8. A study by Borges et al. (2008) found that boarfish may have accounted for as much as $5 \%$ of the total catch of Dutch
pelagic freezer trawlers. Boarfish are also discarded in whitefish fisheries, particularly by Spanish demersal trawlers (Table 3.1.2.3).

### 3.1.3 The fishery in 2017

In 2017 a total of 17388 t of boarfish were caught (Table 3.1.2.1). Ireland continued to be the main participant landing 15484 t but is almost $20 \%$ below its 18858 quota. Denmark landed only 548 t , not even $10 \%$ of its national quota of 6696 t . UK reported almost null boarfish landings. Discards accounted for 1173 tonnes overall. Table 3.1.2.2 shows that about $90 \%$ of the Irish landings were taken in ICES divisions 7.h and 8.a. Thirty-five Irish registered fishing vessels reported catches with almost the entirety made in Q1 (8570 t) and Q4 (6 270 t ).

### 3.1.4 Regulations and their effects

In 2010, the fishery finished early when the European Commission notified member states that mesh sizes of less than 100 mm were illegal. However, in 2011, the European Parliament voted to change Regulation 850/1998 to allow fishing for boarfish using mesh sizes ranging from 32 to 54 mm . The TAC ( 33000 t ) that was introduced in 2011 significantly reduced landings.

### 3.1.5 Changes in fishing technology and fishing patterns

The expansion of the fishery in the mid-2000s was associated with developments in the pumping and processing technology for boarfish catches. These changes made it easier to pump boarfish ashore. Efforts are underway to develop a human consumption market and fishery for boarfish. To date the majority of boarfish landings by Danish, Irish and Scottish vessels have been made into Skagen, Denmark and Fuglafjorour, Faroe Islands to be processed into fishmeal. A small number of Irish vessels have landed into Killybegs and Castletownbere, Ireland. These landings into Irish ports were expected to increase in the future with the development of a human consumption fishery but this now seems unlikely.

### 3.1.6 Discards

Since 2003, the major sources of discards are the Dutch pelagic freezer trawlers and both the Irish and Spanish demersal fleets. More sporadic discards are observed in German pelagic freezer trawlers and the UK demersal fleet. In 2016, Lithuania declared discards for the first time. Discard estimates are not obtained from French freezer trawlers, though discard patterns in these fleets are likely to be similar to the Dutch fleet. Discard data from the Portuguese bottom otter trawl fleet in ICES Division 9.a are also available but are not included in the assessment as they are outside the TAC area. Table 3.1.2.3 shows available data.

It is to be expected that discarding occurred before 2003, in demersal fisheries, however it is difficult to predict what the levels may have been.

Discard data were included in the calculation of catch numbers at age. All discards were raised as one metier using the same age length keys and sampling information as for the landed catches. In the absence of better sampling information on discards, this was considered the best approach. This placed the stock in Category A2 for the ICES Advice in October 2013: Discards 'topped up' onto landings calculations. With the introduction of the discard ban in 2015 this stock was placed in A4: Discards known, with discard ban in place in year +1 . As such the advice will be given for catch in ICES Advice October 2014 and onwards.

### 3.2 Biological composition of the catch

### 3.2.1 Catches in numbers-at-age

Catch number-at-age were prepared for Irish, Danish, Dutch, German and English landings using the ALK in Table 3.2.1.1 together with available samples from the fishery (Table 3.2.1.2). This general ALK was constructed based on 814 aged fish from Irish, Danish and Scottish caught samples from 2012 (see the stock annex for a description of ALKs prior to 2012). In 2017 allocations to unsampled metiers were made according to Table 3.2.1.3. In total 14 Irish and 4 Danish samples with the appropriate .5 cm length bin measurements were collected in 2017 (Table 3.2.1.4). These samples covered only the 4 most heavily fished areas out of a total of 16 (Table 3.2.1.5) and equated to one sample per 966 t landed. The samples comprised 1440 fish measured for length frequency.

The results of the application of the ALK to commercial length-frequency data available for the years 2007-2017 to produce a proxy catch numbers-at-age are available in Table 3.2.1.6. Many old fish are still present in catches, though there appears to be a reduction of older ages since 2007. There have been no strong year classes with poor cohort tracking in the catch numbers. A high number of 2 year old are present in the 2015 data but this does not echo in the number of 3 year old fish in 2016. The modal age from 2007-2011 was 6 and in 2012-2017 it was 7. It should be noted that in WGWIDE 2011 and 2012 the +group for boarfish was 20+. This was reduced to 15+ in WGWIDE 2013 due to potential inaccuracy of the age readings of older fish. Ageing was based on the method that has been validated for ages $0-7$ by Hussy et al. (2012a; b). The age range is similar to the published growth information presented by White et al. (2011).

### 3.2.2 Quality of catch and biological data

Table 3.2.1.3 shows allocations that were made to un-sampled metiers in 2017. Lengthfrequencies of the international commercial landings by year are presented in Table 3.2.2.1.

Sampling in the early years of the fishery (2006-2009) was sparse as there was no dedicated sampling programme in place. The sampling programme was initiated in 2010 and good coverage of the landings has been achieved since then. Full details of the sampling programme in the earlier years are presented in the stock annex. Until 2017, boarfish was not included on the DCF list of species for sampling. Irish sampling comprises only samples from Irish registered vessels. Samples are collected onboard directly from the fish pump during fishing operations and are frozen until returning to port, which ensures high quality samples. Each sample consists of approximately 6 kg of boarfish. This equates to approximately 150 fish which, given the limited size range of boarfish, is sufficient for determining a representative length frequency. The established sampling target is one sample per 1000 t of landings per ICES Division, which is also standard in other pelagic fisheries such as mackerel. Since 2017, all fish in each sample should be measured to the 0.5 cm below for length frequency. Following standard protocols 5 fish per 0.5 cm length class should be randomly selected from each sample for biological data collection i.e. otolith extraction, measurement to the 1 mm below and sex and maturity determination.

There is no sampling programme in place for Scottish catches.
The current surplus production model used to assess boarfish is considered an interim measure prior to the development of an aged-based assessment. In 2017, boarfish was included in the list of species to be sampled by the DCMAP which should provide
estimates of catch at age and facilitate the future development of an age-based stock assessment method.

### 3.3 Fishery Independent Information

### 3.3.1 Acoustic Surveys

A full description of the Boarfish Acoustic Survey (BFAS) which was initiated in July 2011 is given in the stock annex. This survey is run in conjunction with the Malin Shelf herring survey. These surveys are collectively known as the Western European Shelf Pelagic Acoustic Survey (WESPAS).

## Change in abundance calculation method

Acoustic data collected during the WESPAS survey since 2016 were analysed using the StoX software package (ICES 2015a). This package was adopted for WGIPS coordinated surveys in 2016 and has been implemented for all international multi-vessel coordinated surveys within the group (IBWSS, IESSNS, IESSNS and HERAS). The Irish Marine Institute has adopted StoX as the primary abundance calculation tool for national and international acoustic survey data going forward as part of a transitional process initiated during WKEVAL (ICES 2015b). A detailed comparative review of the Irish national method and StoX was carried out on herring during WGIPS 2016 using HERAS and IBWSS data. A difference of $1 \%$ in the total herring biomass estimated by the national method compared to the StoX method for HERAS data was found. Abundances at age showed a greater difference which maybe more related to survey design for the 2015 data set. Regardless, the national abundance by age estimates were all contained within the uncertainty levels surrounding the StoX estimates (ICES 2016). The Irish national abundance is thus considered comparable with StoX going forward.

A description of the StoX application can be found at the following weblink: http://www.imr.no/forskning/prosjekter/stox/nb-no. Survey design and execution for the WESPAS survey adhere to guidelines laid out in the Manual for International Pelagic Surveys (IPS) (ICES 2015a).

## Survey results 2018

The estimate of boarfish biomass from 2011 to 2018 is presented in Table 3.3.1.1 and the spatial distribution of the echotraces attributed to boarfish each year can be seen in Figure 3.3.1.1. In 2018, The WESPAS survey was carried out over a 42 day period beginning on the 09 June in the south $\left(47^{\circ} \mathrm{N}\right)$ and working northwards to $59^{\circ} \mathrm{N}$ ending on 24 July. The survey direction was changed in 2017 from south to north to force containment in the southern area by aligning ourselves with the PELGAS survey. Spatial and temporal alignment has much improved with this move and the survey will be continued in this way in years to come. Overall the WESPAS survey provided continuous coverage from $47 \mathrm{~N}^{\circ}$ to $59 \mathrm{~N}^{\circ}$ over 42 days covering relating to an area coverage of almost 56, 403 nmi 2 (boarfish strata) and transect mileage of over 5,200 nmi. In total 42 trawl stations were undertaken with 14 hauls containing boarfish providing 4,807 individual lengths, 2,234 weights and 945 otoliths for use during the analysis.

The 2018 estimate of biomass is 44,000t lower than observed in 2017 (230,000t in 2017, $186,000 t$ in 2018). The low estimate in 2016 ( $70,000 \mathrm{t}$ ) appears to be an outlier. Containment issues in 2016 were addressed and the survey has been conducted from south to north since 2017. The changes were implemented to increase the precision of the survey overall. Approximately $45 \%$ of the stock was observed in the southern survey area
(Celtic Sea, including Celtic Sea Deep and NW Bank areas). Boarfish were found further north than in previous years.
The age composition of the stock in 2018 is dominated by older age classes (> 7 years) with a peak at 10 year old fish. A second peak at $15+$ years appears to be less in 2018 than in previous years. The numbers at age are variable across years, which may be a result of the fact that an age at length key is used.
The BFAS component of the WESPAS survey is still under development and adaptations have been necessary in an attempt to provide adequate coverage for these species. The survey currently provides an index for both the boarfish and Malin Shelf herring assessments, and in the future, this survey may provide a tuning index for western horse mackerel also. With this in mind, compromises are necessary. A visual comparison of boarfish distribution between years Figure 3.3.1.1 suggests that stock containment to the east, in the Celtic Sea shelf, was achieved in 2014 and possibly 2011 only.

### 3.3.2 International bottom trawl survey (IBTS) Indices Investigation

The western IBTS data and CEFAS English Celtic Sea Groundfish Survey were investigated for their use as abundance indices for boarfish for the first time in 2012. An index of abundance was constructed from the following surveys:

- EVHOE, French Celtic Sea and Biscay Survey, (Q4) 1997 to 2011
- IGFS, Irish Groundfish Survey, (Q4) 2003 to 2011
- WCSGFS, West of Scotland, (Q1 and Q4) 1986 to 2009 (survey design changed in 2010)
- SPPGFS, Spanish Porcupine Bank Survey, (Q3) 2001 to 2011
- SPNGFS, Spanish North Coast Survey, (Q3/Q4) 1991 to 2011
- ECSGFS, CEFAS English Celtic Sea Groundfish Survey, (Q4) 1982 to 2003

From the IBTS data, CPUE was computed as the number of boarfish per 30 min haul. The abundance of boarfish per year per ICES Rectangle (used for visualisation only) was then calculated by summing the boarfish in a given rectangle and dividing by the total number of hauls in that rectangle. Length frequencies are presented in Table 3.3.2.1 for each survey. These surveys cover the majority of the observed range of boarfish in the ICES Area (Figure 3.1). Figure 3.3.2.1 also includes the spatial range of the Portuguese Groundfish Survey (1990-2011), however this survey is outside the current EC TAC area and was never in the assessment.

A detailed analysis of the IBTS data was carried out in 2012 to investigate the main areas of abundance of boarfish in these surveys. This analysis included GAM modelling based on the probability of occurrence of boarfish. The full details of this work are presented in the stock annex. The IBTS appears to give a relative index of abundance, with good resolution between periods of high and low abundance. The main centres of abundance in the survey Figure 3.3.2.2 correspond to the main fishing grounds (Figure 3.1.2.1). Figure 3.3.2.3 shows the signal in abundance, increasing in the 1990s, declining again in the early 2000s, before increasing again.

For subsequent surplus production modelling (see Section 3.6.3), biomass indices were extracted from each of the IBTS surveys using a delta-lognormal model (Stefánsson 1996). Many of the surveys exhibited a large proportion of zero tows with occasionally very large tows, hence the decision to explicitly model the probability of a non-zero tow and the mean of the positive tows. A delta-lognormal fit comprises fitting two generalized linear models (GLMs). The first model (binomial GLM) is used to obtain
the proportion of non-zero tows and is fit to the data coded as 1 or 0 if the tow contained a positive or zero CPUE, respectively. The second model is fit to the positive only CPUE data using a lognormal GLM. Both GLMs were fit using ICES rectangle and year as explanatory factor variables. Where the number of tows per rectangle was less than 5 over the entire series, they are grouped into an "others" rectangle. An index per rectangle and year is constructed, according to Stefánsson (1996), by the product of the estimated probability of a positive tow times the mean of the positive tows. The station indices are aggregated by taking estimated average across all rectangles within a year. To propagate the uncertainty, all survey index analyses were conducted in a Bayesian framework using MCMC sampling (Kery 2010). As WinBugs is no longer updated, the analyses were migrated from WinBUGS to JAGS in 2017. Indeed, JAGS has an almost identical language to WinBUGS and its outputs have been proven equivalent to the previous software (Plummer 2003; Spiegelhalter et al. 2003). In 2018, the assessment was reverted back to WinBUGS as it MCMC sampler appeared more efficient than that of JAGS. Still, the outputs derived from both software are highly similar.

### 3.4 Mean weights-at-age, maturity-at-age and natural mortality

Mean weight-at-age was obtained from the ageing studies of Hüssy et al. (2012b). These mean weights are presented in the text table below. The variation in weight-at-age is due to small sample size and seasonal variation in weight and maturity stage.

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MW <br> $(\mathrm{g})$ | 0.84 | 6.65 | 14.6 | 19.5 | 23.7 | 26.8 | 33.3 | 37.7 | 40 | 47.1 | 50.2 | 51.2 | 62.8 | 56.4 | 62.2 |
| Age | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 |
| MW <br> $(\mathrm{g})$ | 68.9 | 50.5 | 86.7 | 77.9 | 64.6 | 63.5 | 75 | 86 | 71 | 77 | 84.4 | 79.4 | - | 67.6 | 52.8 |

Maturity-at-age was obtained from the ageing studies of Hüssy et al. (2012a; b) and the reproductive study by Farrell et al. (2012).

| AGE | 0 | 1 | 2 | 3 | 4 | 5 | $6+$ |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| PROP MATURE | 0 | 0 | 0.07 | 0.25 | 0.81 | 0.97 | 1 |

Natural mortality (M) was estimated over the life span of the stock using the method described by King (1995). This method assumes that M is the mortality that will reduce a population to $1 \%$ of its initial size over the lifespan of the stock. Based on a maximum age of $31, \mathrm{M}$ is calculated as follows

$$
M=-\ln (0.01) / 31
$$

Following this procedure $M=0.16$ year $^{-1} . M=0.16$ is considered a good estimate of natural mortality over the life span of this boarfish stock, as it is similar to the total mortality estimate from 2007, $(Z=0.18$, see Section 3.6.5). Given that catches in 2007 were relatively low, this estimate of total mortality is considered a good estimate of natural mortality, assuming negligible fishing mortality in previous years.

Similarly, total mortality was estimated from age-structured IBTS data from 2003 to 2006 (years from which data was available for all areas). The total mortality is considered a good estimate of natural mortality as fishing mortality was assumed to be negligible during this period. Total mortality ranged from $0.09-0.2$ with a mean of 0.16 .

The special review in 2012, questioned the validity of a single estimate of $M$ across the entire age range. If an age based assessment is possible in the future, age specific estimates of natural mortality are required. However, the current estimate of M, which covers the whole age range, is considered appropriate in the context of the current situation where age data are used as an indicator approach, rather than as a full assessment method. Given that Z and F are also calculated over the entire (fully selected) range (Section 3.6.5) a single value of M is considered appropriate.

### 3.5 Recruitment

The IBTS data were explored as indices of abundance of 1 year old, and 1-5 years old as a composite recruitment index (Figures 3.5.1 \& 3.5.2). The EVHOE and SPNGFS surveys provide the best indices of recruitment as this is where the juveniles appear to be most abundant (Table 3.3.2.1). It appears that recruitment was high in the late 1990s but declined to a low in 2003. However, this apparent dip in recruitment was not observed in the commercial catch-at-age data. The recruitment signal for ages 1-5 combined has been stable since 2004 with a small increase evident in 2015. The recruitment signal for 1 year old shows a more variable pattern with an increase in 2015 also evident (Figure 3.2.1.1). In 2016, almost all values for age 1 and combined ages 1-5 decreased compared to 2015. The decreases were rather important in the SPNGFS survey and led to historical lows for this survey.

### 3.6 Exploratory assessment

In 2012, a new stock assessment method for Boarfish was tested. In 2013 this Bayesian state space surplus production model (BSP; Meyer \& Millar (1999)) was further developed following reviewers' recommendations in 2012. Different applications of a Bayesian biomass dynamic model were run in 2013 incorporating combinations of catch data, abundance data from the groundfish surveys, and estimates of biomass (and associated uncertainty) from the acoustic surveys (see stock annex for more details of the sensitivity runs). The model and settings from the final accepted run in 2013 were used as the basis of ICES category 1 advice for catch in 2014. However, in 2014 there was concern about the use of the production model for a number of reasons and ICES considered this model as no longer suitable for providing category 1 advice. Since 2014, the assessment model has been used as a basis for trends for providing DLS advice (ICES category 3). ICES considers the current basis for the advice on this stock to be an interim measure prior to development of an age-based assessment.

### 3.6.1 IBTS data

The common ALK (Table 3.2.1.1) was applied to the IBTS number-at-length data. The length-frequency is presented in Table 3.3.2.1and the age-structured index in Table 3.6.1.1 and Figure 3.6.1.1. A cohort effect can be seen with those cohorts from the early 2000s appearing weak. This coincides with a decline in overall abundance in the early 2000s. From the mid-2000s onwards recruitment improved as observed in the abundance of 1-5 year olds in the EVHOE and Spanish northern shelf surveys (Figures 3.5.1 $\& 3.5 .2$ ). It should be noted however that the IBTS data is measured to the 1.0 cm not the 0.5 cm until 2015. Therefore, application of the common ALK to this data must be viewed with caution.

Some of the IBTS CPUE indices displayed marked variability with a large proportion of zero tows and occasionally very large tows (e.g. West of Scotland survey, Figure B.4.7 stock annex). More southern surveys displayed a consistently higher proportion
of positive tows. The variability of the data is reflected in the estimated mean CPUE indices (Figure 3.6.1.2). The West of Scotland survey index had been increasing between 2000 and 2009 but is uncertain, whereas the estimated indices from the other series are typically less variable (Figure 3.6.1.2). In 2014 four of the five current bottom trawl surveys experienced a sharp decline in CPUE, particularly the West of Scotland, the Spanish North Coast, the Spanish Porcupine and Irish Groundfish surveys. Both Spanish surveys remained low in 2015 whereas the latest IGFS and EVHOE surveys indicate an increase. In 2016, values were similar to those of the previous year for all surveys. In 2017, surveys suggest that the stock abundance increased compared to the year before. The only exception is the EVHOE survey but its coverage was only partial year due its research vessel breakdown. The CEFAS English Celtic Sea Groundfish Survey displays a steady increase from the mid-1980s to 2002 with a large but somewhat uncertain estimate in 2003 (Figures 3.6.1.2 \& 3.6.1.3). The spatial extent of each survey is shown in Figure 3.3.2.1.

Diagnostics from the positive component of the delta-lognormal fits indicate relatively good agreement with a normal distribution on the natural logarithmic scale (Figure 3.6.1.4). There is an indication of longer tails in some of the surveys (e.g. WCSGFS, SPPGFS).

Pair-wise correlation between the annual mean survey indices varied. The IGFS, EVHOE and SPNGFS displayed positive correlation (Figure 3.6.1.5). The WCSGFS also displayed a negative correlation with the 2 Spanish surveys (SPPGFS and SPNGFS). The SPPGFS also displayed a negative correlations with EVHOE (Figure 3.6.1.5). Weighting the correlations by the sum of the pair-wise variances resulted in a largely similar correlation structure, though the WCSGFS and SPPGFS were more strongly correlated with the ECSGFS (Figure 3.6.1.6). Note that though some surveys displayed weak or no correlation, we did not a-priori exclude any surveys from the assessment. Sensitivity tests were conducted in 2013, which led to the exclusion of the surveys mentioned previously (see the stock annex).

### 3.6.2 Biomass estimates from acoustic surveys

The Boarfish Acoustic Survey (BFAS) series was initiated in July 2011 and is now in its 8 year. The initial survey in 2011 collected data over 24 hours. Since 2012, acoustic data has been collected between the hours of 04:00 and 00:00. The 2011 data was reworked in 2015 to exclude the data between 00:00 and 04:00. A TS model of -66.2 dB was developed in 2013 [Fässler et al. (2013); odonnell_implementation_2013] and is applied to all surveys in the time series (Figure 3.3.1.1). Over the time series of the survey total biomass has been estimated in the range 863 kt (in 2012) to $70 \mathrm{kt}(2016)$. The precision on the estimates has been good, with coefficients of variation in the range 11 to 21 . An overall downward trend is evident in the first years while estimates have been more stables and rather low since 2014. No strong evidence exists for removing any of the survey points from the time series although 2016 may look like an outlier.

It should be noted that two acoustic surveys are conducted annually to the south of the southern limit of the dedicated Boarfish survey. In 2016 the PELACUS recorded an increase in biomass from 2015 although not of the order of the decrease seen further north. The Spanish PELGAS surveys recorded low levels of biomass, similar to that in 2015. Both these surveys take place 2-3 months prior to the boarfish survey.

### 3.6.3 Biomass dynamic model

In 2012 an exploratory biomass dynamic model was developed. This was a Bayesian state space surplus production model (Meyer \& Millar 1999), incorporating the catch data, IBTS data, and acoustic biomass data. This assessment was then peer-reviewed by two independent experts on behalf of ICES. In 2013 a new assessment was provided, which was based on the previous year's work and the reviewers' comments and formed the basis of a category 1 assessment. Details of the review and the associated changes can be found in the stock annex.

In 2014 the Bayesian state space surplus production model was again fit using the catch data, delta-lognormal estimated IBTS survey indices, and the acoustic survey estimates. However, the inclusion of the low 2014 acoustic biomass estimate changed the perception on the stock, which raised concerns over the sensitivity and process error of the model. The stock was moved from a category 1 assessment to a category 3 with the results of the surplus production model being used to calculate an index for the data limited stock approach.

Since 2014, the procedure used to run the model did not change. Only the length of the time series used increase yearly. Details of this exploratory run used to calculate the DLS index are described below. Further model development work is undertaken since 2015 but did not lead to any change so far.

In the Bayesian state space surplus production model the biomass dynamics are given by a difference form of a Schaefer biomass dynamic model:

$$
B_{t}=B_{t-1}+r B_{t-1}\left(1-\frac{B_{t-1}}{K}\right)+C_{t-1}
$$

where $B_{t}$ is the biomass at time $\mathrm{t}, \mathrm{r}$ is the intrinsic rate of population growth, $K$ is the carrying capacity, and $C_{t}$ is the catch, assumed known exactly. To assist the estimation the biomass is scaled by the carrying capacity, denoting the scaled biomass $P_{t}=B_{t} / \mathrm{K}$. Lognormal error structure is assumed giving the scaled biomass dynamics (process) model:

$$
P_{t}=\left(P_{t-1}+r P_{t-1}\left(1-P_{t-1}\right)+\frac{C_{t-1}}{K}\right) e^{\mu_{t}}
$$

where the logarithm of process deviations are assumed normal $u_{t}=N\left(0, \sigma_{2}^{\mu}\right)$ with $\sigma_{2}^{\mu}$ the process error variance.

The starting year biomass is given by $a K$, where a is the proportion of the carrying capacity in the first year. The biomass dynamics process is related to the observations on the indices through the measurement error equation:

$$
I_{j, t}=q_{j} P_{t} K e^{\varepsilon_{j, t}}
$$

where $I_{j, t}$ is the value of abundance index $j$ in year $t, q_{j}$ is survey-specific catchability, $B_{t}$ $=P_{t} K$, and the measurement errors are assumed lognormally distributed with $u_{t}=$ $N\left(0, \varepsilon_{e, j, t}^{2}\right)$ where $\varepsilon_{e, j, t}^{2}$ is the index-specific measurement error variance. $\operatorname{Var}\left(I_{j, t}\right)$ is obtained from the delta-lognormal survey fits. That is, the variance of the mean annual estimate per survey is inputted directly from the delta-lognormal fits (Figure 3.6.1.2) as opposed to estimating a measurement error within the assessment. The measurement error is obtained from:

$$
\sigma_{e, j, t}^{2}=\ln \left(1+\frac{\operatorname{Var}\left(I_{j, t}\right)}{\left(I_{j, t}\right)^{2}}\right)
$$

For the acoustic survey, the CV of the survey was transformed into a lognormal variance via

$$
\sigma_{\varepsilon, \text { acoustic }, t}^{2}=\ln \left(C V_{\text {acoustic }, t}^{2}+1\right)
$$

Prior assumptions on the parameter distributions were:

- Intrinsic rate of population growth: $r \sim U(0.001,2)$
- Natural logarithm of the carrying capacity: $\ln (K) \sim U(\ln (\max (C)$, $\ln (10 . \operatorname{sum}(C))=U(\ln (144047), \ln (4450407))$
- Proportion of carrying capacity in first year of assessment: $a \sim U[0.001,1.0]$
- Natural logarithm of the survey-specific catchabilities $\ln \left(q_{i}\right) \sim U(-16,0)$ (for IBTS only). The acoustic survey prior is discussed below.
- Process error precision $\frac{1}{\sigma_{u}^{2}} \sim \operatorname{gamma}(0.001,0.001)$


## Specification

During the 2013 WGWIDE meeting a number of different iterations of the model were run to discern the best parameters for the assessment. After four initial runs and four sensitivity runs the settings for the final run (run 2.2) were chosen. These settings are shown below and were used for the assessment model since 2014. (More details of the trial runs in 2013 can be found in the stock annex).

The specifications for the final boarfish assessment model runs are:

## Acoustic survey

Years: 2011-2018
Index value (Iacoustic, y): 'total' in tonnes (i.e. Definitely Boarfish + Probably Boarfish + Boarfish in a Mix)

Catchability ( $q_{\text {acoustic }}$ ): A free, but strong prior (i.e. the acoustic survey is treated as a relative index but is strongly informed, this allows the survey to cover $<100 \%$ of the stock).

## IBTS surveys

6 delta log normal indices (WCSGFS, SPPGFS, IGFS, ECSGFS, SPNGFS, EVHOE)
First 5 and last 7 (since 2017, because of change in survey design) years omitted from WCSGFS

## First 9 years omitted from ECSGFS

Following plenary discussion of the sensitivity runs in 2013, it was decided that the final run be based on a run that includes all surveys with the omission of the first 5 years of the WCSGFS and first 9 years of the ECSGFS. The reasons for this decision were: *it is unclear whether boarfish were consistently recorded in the early part of the ECSGFS, * the WCSGFS is thought to be at the northern extreme of the distribution and may not be an appropriate index for the whole stock, * the SPNGFS commences in 1991 such that running the assessment from 1991 onwards includes at least three surveys without relying, solely on the ECSGFS and WCSGFS, * surveys are internally weighted such that highly uncertain values receive lower weight.

## Catches

2003-2018 time series

## Priors

The final run assumes a strong prior $\ln \left(q_{\text {acoustic }}\right) \sim N(1,1 / 4)$ (mean 1 , standard deviation $0.25)$, which has $95 \%$ of the density between 0.5 and 2 . Given the short acoustic series (6 years) it is not possible to estimate this parameter freely (i.e. using an uninformative
prior). The prescription of a strong prior removes the assumption of an absolute index from the acoustic survey. This assumption will be continually updated as additional data accrue.

## Run convergence

Parameters for the 2018 model run converged with good mixing of the chains and Rhat values lower than 1.1 indicating convergence (Figures 3.6.3.1 \& 3.6.3.2). MCMC chain autocorrelation was rather high but was compensated by long MCMC chains providing representative samples of the parameter posteriors (Figure 3.6.3.3).

Diagnostic plots are provided in Figure 3.6.3.4 showing residuals about the model fit. A fairly balanced residual pattern is evident. In some cases outliers are apparent, for instance in the English survey in the final year (2003). However, these points are downweighted according to the inverse of their variance and hence do not contribute much to the model fit. The west of Scotland IBTS survey, located at the northern extreme of the stock distribution underestimates the stock in the early period (years) and overestimates it in the recent period from all fits. This could be indicative of stock expansion into this area at higher stock sizes and suggests that this index is not representative of the whole stock. `Figure 3.6.3.5 shows the prior and posterior distributions of the parameters of the biomass dynamic model. The estimate of $q$ is less than 1.0, leading to a higher estimate of final stock biomass than the acoustic survey.

## Results

Trajectories of observed and expected indices are shown in Figure 3.6.3.6, along with the stock size over time and a harvest ratio (total catch divided by estimated biomass). Parameter estimates from the model run are summarized in Table 3.6.3.1. Biomass in 2018 is estimated to be 284770 t and it appears to be stable but low over the last 5 years. It is worth noting that the extremely low biomass estimate from the 2016 acoustic survey now appears considered as an outlier by the model. As a consequence the 2016 biomass estimate increased from 108000 t last in 2016 to about 240000 t in 2017 and 2018. Retrospective plots of TSB and F, presented in Figure 3.6.3.7, show that the perception of the stock is stable through time with the exception of 2013 prior to the inclusion of the lower biomass estimates of the acoustic surveys since 2014.

### 3.6.4 Pseudo-cohort analysis

Pseudo-cohort analysis is a procedure where mortality is calculated by means of catch curves derived from catch-at-age from a single year. This is in contrast to cohort analysis, which is the basis of VPA-type assessments. In cohort analysis, mortality is calculated across the ages of a year class, not within a single year. Because only seven years of sampling data were available and owing to the large age range currently in the catches a cohort analysis would only yield information for a very limited age and year range. Therefore, pseudo-cohort analysis was performed to supplement the Bayesian state space model.

Pseudo-cohort $Z$ estimates increased with the rapid expansion of the fishery but decreased in 2011 due to the introduction of the first boarfish TAC (Table 3.6.4.1). By subtracting $M(=0.16)$, an estimate of $F$ was obtained for each year (ages $7-14$ ). This series was revised to represent ages 7-14, rather than 6-14 as in previous years, because in 2013 age 6 boarfish were not fully selected, i.e. age 7 had higher abundance at age.

It can be seen from the text table below that $Z=M$ in 2007, the initial year of the expanded fishery, while $F$ is negligible. $F$ increased to a high of 0.29 in 2012 and has gradually reduced down to 0.15 in 2015 and 2016. In 2017, it increased up to 0.17. There was
a weak correlation between catches and pseudo-cohort $F\left(r^{2}=0.48\right)$. Recent $F$ estimated this way is close to $F M S Y$ ( 0.149 ) and above $F 0.1$ (0.13).

| YEAR | Z.(7-14) | F.(Z-M) | CATCH.(T) |
| :---: | :---: | :---: | :---: |
| 2007 | 0.17 | 0.01 | 21576 |
| 2008 | 0.33 | 0.17 | 34751 |
| 2009 | 0.36 | 0.2 | 90370 |
| 2010 | 0.33 | 0.17 | 144047 |
| 2011 | 0.29 | 0.13 | 37096 |
| 2012 | 0.45 | 0.29 | 87355 |
| 2013 | 0.36 | 0.2 | 75409 |
| 2014 | 0.37 | 0.21 | 45231 |
| 2015 | 0.31 | 0.15 | 17766 |
| 2016 | 0.31 | 0.15 | 19315 |
| 2017 | 0.33 | 0.17 | 17388 |

### 3.6.5 State of the stock

According to this year assessment, total stock biomass appeared to increase from a low to average level from the early to mid-1990s (Figure 3.6.3.6). The stock fluctuated around this level until 2009, when it increased until 2012, followed by a sharp decline from 2013 to 2014. Since 2014, the abundance appear low but rather stable, fluctuating around 320000 t . There was concern in 2014 that this decline was exaggerated by an unusually low acoustic biomass estimate that led to a downward revision in stock trajectory. However, the 2014 survey may now be viewed as one of the most successful in terms of containment. The comparably low 2014 biomass estimate was supported by results of the 2015 survey. The 2016 biomass estimate, the lowest of the time series now appears as an outlier and do no longer drive the stock abundance estimates to even lower values. The uncertainty surrounding the estimates of biomass the last years remain important with wide $95 \%$ credible interval (Table 3.6.5.1). This reflects the uncertainty in the survey indices, and short exploitation history of the stock and the treatment of the acoustic survey as a relative biomass index. As more data accumulates from this survey, it is expected that the prior will become increasingly updated, and potentially less variable.

Catch data are available from 2001, the first year of commercial landings, and reasonably comprehensive discard data are available from 2003. Peak catches were recorded in 2010, when over 140000 t were taken. Elevated fishing mortality was observed, associated with the highest recorded catch in 2010. Fishing mortality, expressed as a harvest ratio (catch divided by total biomass), was first recorded in 2003. Before that time, it is to be expected that some discarding took place, and there were some commercial landings. Fishing mortality increased measurably from 2006, reaching a peak in 20092010. F declined in 2011 as catches became regulated by the precautionary TAC but increased year on year until 2015 when reduced catches resulted in a reduction. The considerable catches in recent years do not appear to have significantly truncated the size or age structure of the stock and $15+$ group fish are still abundant (Figure 3.2.1.1).

Since 2017, MSY reference points have been developed for the boarfish stock and may be used to guide the advice. The ICES MSY framework specifies a target fishing mortality, FMSY (stock growth rate over 2), which, over the long term, maximises yield, and also a spawning biomass, MSY B trigger (stock carrying capacity over 4), below which target fishing mortality should be reduced linearly relative to the SSB B trigger $^{\text {ratio. In }}$ 2018, FMSY and MSY Btrigger are estimated respectively equal to 0.185 (parameter r / 2)
and $165420 t$ (parameter K / 4). Throughout the history of the fishery, estimates of stock biomass have remained above MSYB trigger. $_{\text {. Fishing mortality }}(\mathrm{F}$ ) was greater than FMSY in 2009, 2010 and 2014, but has decreased since. In 2018, the stock is in the green area of the Kobe plot (Figure 3.6.6.1).

Estimates of recruitment are not available from the stock assessment. However, an independent index of recruitment is available from groundfish surveys (Section 3.5). Observations from the survey recruitment of 1 year olds show strong negative trends since 2010 (Figure 3.5.1) and a weaker, but still negative, trend for ages 1-5 combined (Figure 3.5.2) for 2 out of 3 surveys. The trend within the IGFS is opposite.

### 3.7 Short Term Projections

As the assessment is exploratory, no short term projections were conducted.

### 3.8 Long term simulations

No long term simulations were conducted.

### 3.9 Candidate precautionary and yield based reference points

### 3.9.1 Yield per Recruit

A yield per recruit analysis was conducted in 2011 (Minto et al. 2011) and F0.1 was estimated to be 0.13 whilst $\mathrm{F}_{\text {max }}$ was estimated in the range 0.23 to 0.33 (Figure 3.9.1.1). F0.1 was considered to be well estimated (Figure 3.9.1.2). No new yield per recruit analyses were performed in subsequent years.

### 3.9.2 Precautionary reference points

It does not appear that boarfish is an important prey species in the NE Atlantic (Section 3.13). ICES (2007) considered that precautionary $F$ targets ( $F p a$ ) should be consistent with F130 625 t based on the exploratory assessment in 2018).

### 3.9.3 Other yield based reference points

Yield per recruit analysis, following the method of Beverton \& Holt (1957), found F0.1 to be robustly estimated at 0.13 (ICES 2011; Minto et al. 2011).

### 3.10 Quality of the assessment

ICES considers the current basis for the advice on this stock to be an interim measure prior to development of an age-based assessment. In addition, the acoustic survey used (BFAS / WESPAS) is in a state of development at present and there are concerns that the acoustic survey may not be containing the stock sufficiently. The assessment was downgraded from Category 1 to Category 3 in 2014, and it has remained in this category since. The model is still considered suitable for category 3 advice, because it provides the best means of combining the available survey series. The assessment is very sensitive to the acoustic series. In addition, a substantial part of the year to year variations in the stock abundance is linked to the process error. The use of some priors (like ratio to virgin biomass in the first year of the assessment) and survey (WCSGFS for instance) may need to be revised.

Additional work to improve the surplus production model is undertaken since 2015 and will continue next year. A issue list has been provided and a benchmark is planned for 2020.

The bottom trawl survey data are considered to be a good index of abundance given that boarfish aggregate near the bottom at this time of year. The trawl surveys record high abundances of the species, but with many zero hauls. The delta-lognormal error structure used in the analyses is considered to be a good means of dealing with such data. The biomass dynamic model used in the stock assessment is based on the recent benchmarked assessment of megrim in Sub-divisions 4 and 6. The model was further developed by including acoustic survey biomass estimates. One drawback of the model is that it does not provide estimates of recruitment. However, an estimate of recruitment strength is available from the Spanish and French trawl surveys.

### 3.11 Management considerations

As this stock is now placed in category 3, the ICES advice for 2018 is based on harvest control rules for data limited stocks (ICES 2017). Since the biomass estimate from the Bayesian model is considered reliable for trend based assessment, an index can be calculated according to Method 3.1 of ICES (2012). The advice is based on a comparison of the average of the two most recent index values with the average of the three preceding values multiplied by the most recent catch. Table 3.6.5.1 shows the biomass estimates from the model from which the index was calculated.

ADG decided to use the advice given in 2017 and based on this framework for 2 years. This results in an advised catch of 21830 t for 2019. More details can be found in last year report. The apparent stability of the assessment this year comforts this decision.

Although no longer accepted as the basis for an analytic assessment, the surplus production model still provides the best unified view of this stock (Figure 3.6.3.6).

### 3.12 Stock structure

A dedicated study on the stock structure of boarfish within the Northeast Atlantic and Mediterranean Sea commenced in October 2013 in order to resolve outstanding questions regarding the stock structure of boarfish and the suitability of assessment data. Results (Farrell et al. 2016) indicated strong population structure across the distribution range of boarfish with 7-8 genetic populations identified (Figure 3.12.1).

The eastern Mediterranean (MED) samples comprised a single population and were distinct from all other samples. Similarly the Azorean (AZA), Western Saharan (MOR) and Alboran ( $A L M$ ) samples were distinct from all others. Of particular relevance to the assessment and management of the boarfish fishery is the identification and delineation of the population structure between southern Portuguese waters (PTN2B-PTS) and waters to the geographic north. A distinct and temporally stable mixing zone was evident in the waters around Cabo da Roca. The PTN2A sample appeared to be significantly different from all other samples however this sample was relatively small and was considered to represent a mixed sample rather than a true population.

No significant spatial or temporal population structure was found within the samples comprising the NEA population (Figure 3.12.1). A statistically significant but comparatively low level of genetic differentiation was found between this population and the northern Spanish shelf/northern Portuguese samples (NSA-PTN1). However, a high level of migration was revealed between these two populations and no barriers to gene flow were detected between them. Therefore, for the purposes of assessment and management these areas can be considered as one unit.

Analyses indicated a lack of significant immigration into this northeast Atlantic boarfish stock from populations to the south or from insular elements and the strong genetic differentiation among these regions indicate that the purported increases in abundance in the northeast Atlantic area are not the result of a recent influx from other regions. The increase in abundance is most likely the result of demographic processes within the northeast Atlantic stock (Blanchard \& Vandermeirsch 2005; Coad et al. 2014).
Whilst the current assessment and management area constitutes the majority of the most northern population it should be extended into Northern Portuguese waters and repeated genetic monitoring of the stock in this region should be conducted to ensure the validity of this delineation. Based on analyses of IBTS data (ICES 2013) the biomass in this area is suspected to be small relative to the overall biomass in the TAC area.

### 3.13 Ecosystem considerations

The ecological role and significance of boarfish in the NE Atlantic is largely unknown. However, in the southeast North Atlantic, in Portuguese waters, they are considered to have an important position in the marine food web (Lopes et al. 2006). The diet has been investigated in the eastern Mediterranean, Portuguese waters and at Great Meteor Seamount and consists primarily of copepods, specifically Calanus helgolandicus, with some mysid shrimp and euphausiids (Macpherson 1979; Fock et al. 2002; Lopes et al. 2006). This contrasted with the morphologically similar species, the slender snipefish, Macroramphosus gracilis and the longspine snipefish, M. scolopax, whose diet comprised Temora spp., copepods and mysid shrimps, respectively (Lopes et al. 2006). Despite the obvious potential for these species to feed on fish eggs and larvae, there was no evidence to support this conclusion in Portuguese waters and they were not considered predators of commercial fishes and thus their increase in abundance was unlikely to affect recruitment of commercial fish species. If the NE Atlantic population of boarfish is sufficiently large then there exists the possibility of competition for food with other widely distributed planktivorous species.
Both seasonal and diurnal variations were observed in the diet of boarfish in all three regions. In the eastern Mediterranean and Portuguese waters, mysids become an important component of the diet in autumn, which correlates with their increased abundance in these regions at this time (Macpherson 1979; Lopes et al. 2006). Fock et al. (2002) found that boarfish at Great Meteor Seamount fed mainly on copepods and euphausiids diurnally and on decapods nocturnally, indicating habitat dependent resource utilization.

Boarfish appear an unlikely target of predation given their array of strong dorsal and anal fin spines and covering of ctenoid scales. However, there is evidence to suggest that they may be an important component of some species' diets. Most studies have focused in the Azores and few have mentioned the NE Atlantic, probably due to the relatively low abundance in the region until recent years. In the Azores, boarfish was found to be one of the most important prey items for tope (Galeorhinus galeus), thornback ray (Raja clavata), conger eel (Conger conger), forkbeard (Phycis phycis), bigeye tuna (Thunnus obesus), yellowmouth barracuda (Sphyraena viridensis), swordfish (Xiphias gladius), blackspot seabream (Pagellus bogaraveo), axillary seabream (Pagellus acarne) and blacktail comber (Serranus atricauda) (Clarke et al. 1995; Morato et al. 1999, 2000, 2001, 2003; Arrizabalaga et al. 2008). Many of these species also occur in the NE Atlantic shelf waters although it is unknown whether boarfish represent a significant component of the diet in this region.

In the NE Atlantic boarfish have not previously been recorded in the diets of tope or thornback ray (Holden \& Tucker 1974; Ellis et al. 1996). However, this does not prove that they are currently not a prey item. A study of conger eel diet in Irish waters from 1998-1999 failed to find boarfish in the $\operatorname{diet}(\mathrm{O} \backslash \& \backslash \# 39$ et al. 2004). However, in Portuguese waters a recent study has found boarfish to be the most numerous species in the diet of conger eels (Xavier et al. 2010). It has been suggested that boarfish are an important component of the diet of hake (Merluccius merluccius), as they are sometimes caught together. However, a recent study of the diet of hake in the Celtic Sea and Bay of Biscay did not report any boarfish in the stomachs of hake caught during the 2001 EVHOE survey (Mahe et al. 2007).

The conspicuous presence of boarfish in the diet of so many fish species in the Azores is perhaps more related to the lack of other available food sources than to the palatability of boarfish themselves. Given the large abundance in NE Atlantic shelf waters it is likely that they would have been recorded more frequently if they were a significant and important prey item.

Boarfish are also an important component of the diet a number of sea birds in the Azores, most notably the common tern (Sterna hirundo, Granadeiro et al. (2002)) and Cory's shearwater (Calonectris diomedea, Granadeiro et al. (1998)). This is surprising given that in the Mediterranean discarded boarfish were rejected by seabirds whereas in the Azores they were actively preyed on (Oro \& Ruiz 1997). Cory's shearwaters are capable of diving up to 15 m whilst the common tern is a plunge-diver and may only reach 2-3 m . It is therefore surprising that boarfish are such a significant component of their diet given that it is generally considered a deeper water fish. In the Azores boarfish shoals are sometimes driven to the surface by horse mackerel and barracuda where they are also attacked by diving sea birds (J. Hart, CW Azores, pers. comm.). Anecdotal reports from the Irish fishery indicate that boarfish are rarely found in waters shallower than 40 m . This may suggest that they are outside the range of shearwaters and gannets, the latter having a mean diving depth of $19.7 \pm 7.5 \mathrm{~m}$ (Brierley \& Fernandes 2001). However, the upper depth range of boarfish is within maximum diving depth recorded for auks ( 50 m ) as recorded by Barrett \& Furness (1990). Given their frequency in the diets of marine and bird life in the Azores, boarfish appear to be an important component of the marine ecosystem in that region. There is currently insufficient evidence to draw similar conclusions in the NE Atlantic.

The length-frequency distribution of boarfish may be important to consider. IBTS data shows an increase in mean total length with latitude Table 3.3.2.1 and perhaps the smaller boarfish in the southern regions are more easily preyed upon. Length data of boarfish from stomach contents studies of both fish and sea birds in the Azores indicate that the boarfish found are generally $<10 \mathrm{~cm}$ (Granadeiro et al. 1998, 2002).

### 3.14 Proposed management plan

In 2015 the Pelagic Advisory Council submitted a revised draft management strategy for Northeast Atlantic boarfish. The EU has requested ICES to evaluate the following management plan:

This management strategy aims to achieve sustainable exploitation of boarfish in line with the precautionary approach to fisheries management, FAO guidelines for new and developing fisheries, and the ICES form of advice.

1 ) The TAC shall be set in accordance with the following procedure, depending on the ICES advice
a) If category 1 advice (stocks with quantitative assessments) is given based on a benchmarked assessment, the TAC shall be set following that advice.
b ) If category 1 or 2 (qualitative assessments and forecasts) advice is given based on a non-benchmarked assessment the TAC shall be set following this advice.
c ) Categories 3-6 are described below as follows:
i) Category 3: stocks for which survey-based assessments indicate trends. This category includes stocks with quantitative assessments and forecasts which for a variety of reasons are considered indicative of trends in fishing mortality, recruitment, and biomass.
ii ) Category 4: stocks for which only reliable catch data are available.This category includes stocks for which a time series of catch can be used to approximate MSY.
iii) Category 5: landings only stocks. This category includes stocks for which only landings data are available.
2 ) Category 6: Category 6 - negligible landings stocks and stocks caught in minor amounts as bycatch
3 ) Notwithstanding paragraph 1, if, in the opinion of ICES, the stock is at risk of recruitment impairment, a TAC may be set at a lower level.
4 ) If the stock, estimated in the either of the 2 years before the TAC is to be set, is at or below $B_{\lim }$ or any suitable proxy thereof, the TAC shall be set at $0 t$.
5 ) The TAC shall not exceed $75,000 t$ in any year.
6 ) The TAC shall not be allowed to increase by more than $25 \%$ per year. However, there shall be no limit on the decrease in TAC.
7 ) Closed seasons, closed areas, and moving on procedures shall apply to all directed boarfish fisheries as follows:
a) A closed season shall operate from 31st March to 31st August. This is because it is known that herring and mackerel are present in these areas and may be caught with boarfish.
b ) A closed area shall be implemented inside the Irish 12-miles limit south of $52^{\circ} 30$ from $12^{\text {th }}$ February to $31^{\text {st }}$ October, in order to prevent catches of Celtic Sea herring, known to form aggregations at these times.
c ) If catches of other species covered by a TAC amount to more than $5 \%$ of the total catch by day by ICES statistical rectangle, then all fishing must cease in that rectangle for 5 consecutive days.

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Table 3.1.2.1. Boarfish in ICES Subareas 27.6, 7, 8. Landings, discards and TAC by country by year (t), 2001-2017. (Data provided by Working Group members). These figures may not in all cases correspond to the official statistics and cannot be used for management purposes.

| $\begin{gathered} \text { Yea } \\ \text { rs } \end{gathered}$ | Denm ark | Germ any | Irela <br> nd | The.Nethe rlands | UK.Eng <br> land | UK.Scot land | Unalloc ated | Disca rds | Tot al | $\begin{aligned} & \mathrm{TA} \\ & \mathrm{C} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 |  |  | 120 |  |  |  |  |  | 120 | - |
| 2002 |  |  | 91 |  |  |  |  |  | 91 | - |
| 2003 |  |  | 458 |  |  |  |  | 10929 | $\begin{aligned} & 113 \\ & 87 \end{aligned}$ | - |
| 2004 |  |  | 675 |  |  |  |  | 4476 | $515$ | - |
| 2005 |  |  | 165 |  |  |  |  | 5795 | $\begin{aligned} & 595 \\ & 9 \end{aligned}$ | - |
| 2006 |  |  | 2772 |  |  |  |  | 4365 | $\begin{aligned} & 713 \\ & 7 \end{aligned}$ | - |
| 2007 |  |  | $\begin{aligned} & 1761 \\ & 5 \end{aligned}$ |  |  | 772 |  | 3189 | $\begin{aligned} & 215 \\ & 76 \end{aligned}$ | - |
| 2008 | 3098 |  | $\begin{aligned} & 2158 \\ & 5 \end{aligned}$ |  |  | 0.45 |  | 10068 | $\begin{aligned} & 347 \\ & 51 \end{aligned}$ | - |
| 2009 | 15059 |  | $\begin{aligned} & 6862 \\ & 9 \end{aligned}$ |  |  |  |  | 6682 | $903$ | - |
| 2010 | 39805 |  | $\begin{aligned} & 8845 \\ & 7 \end{aligned}$ |  |  | 9241 |  | 6544 | $\begin{aligned} & 144 \\ & 04 \end{aligned}$ | - |
| 2011 | 7797 |  | $2068$ |  |  | 2813 |  | 5802 | $370$ | $\begin{aligned} & 330 \\ & 00 \end{aligned}$ |
| 2012 | 19888 |  | $\begin{aligned} & 5594 \\ & 9 \end{aligned}$ |  |  | 4884 |  | 6634 | $873$ | $820$ |
| 2013 | 13182 |  | $\begin{aligned} & 5225 \\ & 0 \end{aligned}$ |  |  | 4380 |  | 5598 | $\begin{aligned} & 754 \\ & 0 \end{aligned}$ | $\begin{aligned} & 820 \\ & 00 \end{aligned}$ |
| 2014 | 8758 |  | $\begin{aligned} & 3462 \\ & 2 \end{aligned}$ |  |  | 38 |  | 1813 | $\begin{aligned} & 452 \\ & 31 \end{aligned}$ | $\begin{aligned} & 133 \\ & 957 \end{aligned}$ |
| 2015 | 29 | 4 | $\begin{aligned} & 1632 \\ & 5 \end{aligned}$ | 375 | 104 |  |  | 929 | $\begin{aligned} & 177 \\ & 66 \end{aligned}$ | $\begin{aligned} & 532 \\ & 96 \end{aligned}$ |
| 2016 | 337 | 7 | $\begin{aligned} & 1749 \\ & 6 \end{aligned}$ | 171 | 21 |  |  | 1284 | $\begin{aligned} & 193 \\ & 15 \end{aligned}$ | $\begin{aligned} & 476 \\ & 37 \end{aligned}$ |
| 2017 | 548 |  | $\begin{aligned} & 1548 \\ & 5 \end{aligned}$ | 182 | 0.13 |  |  | 1173 | $\begin{aligned} & 173 \\ & 88 \\ & \hline \end{aligned}$ | $\begin{aligned} & 272 \\ & 88 \end{aligned}$ |

Table 3.1.2.2. Boarfish in ICES Subareas 27.6, 7, 8. Landings by year (t), 2001-2017 (Data provided by Working Group members). These figures may not in all cases correspond to the official statistics and cannot be used for management purposes.

| Year | Area | Denmark | Germany | Ireland | The.Netherlands | UKE | UKS | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | ALL |  |  | 120 |  |  |  | 120 |
| 2002 | ALL |  |  | 91 |  |  |  | 91 |
| 2003 | ALL |  |  | 458 |  |  |  | 458 |
| 2003 | 6.a |  |  | 65 |  |  |  | 65 |
| 2003 | 7.b |  |  | 214 |  |  |  | 214 |
| 2003 | 7.j |  |  | 179 |  |  |  | 179 |
| 2004 | ALL |  |  | 675 |  |  |  | 675 |
| 2004 | 6.a |  |  | 292 |  |  |  | 292 |
| 2004 | 7.b |  |  | 224 |  |  |  | 224 |
| 2004 | 8.d |  |  | 38 |  |  |  | 38 |
| 2004 | 7.j |  |  | 122 |  |  |  | 122 |
| 2005 | ALL |  |  | 165 |  |  |  | 165 |
| 2005 | 6.a |  |  | 10 |  |  |  | 10 |
| 2005 | 7.b |  |  | 105 |  |  |  | 105 |
| 2005 | 8.a |  |  | 38 |  |  |  | 38 |
| 2005 | 7.j |  |  | 12 |  |  |  | 12 |
| 2006 | ALL |  |  | 2772 |  |  |  | 2772 |
| 2006 | 6.a |  |  | 21 |  |  |  | 21 |
| 2006 | 7.b |  |  | 15 |  |  |  | 15 |
| 2006 | 7.g |  |  | 375 |  |  |  | 375 |
| 2006 | 8.a |  |  | 1 |  |  |  | 1 |
| 2006 | 7.j |  |  | 2360 |  |  |  | 2360 |
| 2007 | ALL |  |  | 17615 |  |  | 772 | 18386 |
| 2007 | 5.b2 |  |  | 6 |  |  |  | 6 |
| 2007 | 6.a |  |  | 93 |  |  |  | 93 |
| 2007 | 7.b |  |  | 1259 |  |  |  | 1259 |
| 2007 | 7.g |  |  | 120 |  |  |  | 120 |
| 2007 | 8.a |  |  | 5 |  |  |  | 5 |
| 2007 | 7.-j |  |  | 16131 |  |  | 772 | 16903 |
| 2008 | ALL |  |  | 21584 |  |  |  | 21585 |
| 2008 | 6.a |  |  | 28 |  |  |  | 28 |
| 2008 | 7.b |  |  | 3 |  |  |  | 3 |
| 2008 | 7.g |  |  | 184 |  |  |  | 184 |
| 2008 | 7.- |  |  | 21370 |  |  |  | 21370 |
| 2009 | ALL |  |  | 68629 |  |  |  | 68629 |
| 2009 | 6.a |  |  | 45 |  |  |  | 45 |
| 2009 | 7.b |  |  | 73 |  |  |  | 73 |
| 2009 | 7.c |  |  | 1 |  |  |  | 1 |
| 2009 | 7.g |  |  | 4912 |  |  |  | 4912 |
| 2009 | 7.h |  |  | 18225 |  |  |  | 18225 |
| 2009 | 7.j |  |  | 45372 |  |  |  | 45372 |


| Year | Area | Denmark | Germany | Ireland | The.Netherlands | UKE | UKS | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | ALL | 39805 |  | 88457 |  |  | 9241 | 137503 |
| 2010 | $6 . a$ |  |  | 1349 |  |  | 10 | 1359 |
| 2010 | 6.aS |  |  | 7 |  |  |  | 7 |
| 2010 | 7.b |  |  | 2258 |  |  |  | 2258 |
| 2010 | 7.c |  |  | 35 |  |  | 4 | 39 |
| 2010 | 7.e | 2 |  |  |  |  |  | 2 |
| 2010 | 7.g | 672 |  | 3649 |  |  |  | 4321 |
| 2010 | 7.h | 1465 |  | 8453 |  |  | 1712 | 11629 |
| 2010 | 7.j | 37667 |  | 72707 |  |  | 7515 | 117889 |
| 2011 | ALL | 7797 |  | 20685 |  |  | 2813 | 31295 |
| 2011 | 6.a |  |  | 26 |  |  |  | 26 |
| 2011 | 7.b |  |  | 274 |  |  |  | 274 |
| 2011 | 7.c |  |  | 9 |  |  |  | 9 |
| 2011 | 7.g |  |  | 811 |  |  |  | 811 |
| 2011 | 7.h | 4155 |  | 8540 |  |  | 2813 | 15508 |
| 2011 | 8.a | 18 |  |  |  |  |  | 18 |
| 2011 | 7.j | 3624 |  | 11025 |  |  |  | 14648 |
| 2012 | ALL | 19888 |  | 55949 |  |  | 4884 | 80720 |
| 2012 | $6 . a$ |  |  | 125 |  |  |  | 125 |
| 2012 | 7.b | 80 |  | 4501 |  |  | 838 | 5419 |
| 2012 | 7.c |  |  | 108 |  |  | 907 | 1015 |
| 2012 | 7.9 |  |  | 616 |  |  |  | 616 |
| 2012 | 7.h | 5837 |  | 10579 |  |  | 3139 | 19554 |
| 2012 | 8.a | 1604 |  | 93 |  |  |  | 1697 |
| 2012 | 7.j | 12366 |  | 39928 |  |  |  | 52294 |
| 2013 | ALL | 13182 |  | 52250 |  |  | 4380 | 69811 |
| 2013 | 6.a |  |  | 538 |  |  | 15 | 553 |
| 2013 | 7.b |  |  | 10405 |  |  | 100 | 10505 |
| 2013 | 7.e |  |  |  |  |  | 883 | 883 |
| 2013 | 7.g |  |  | 1808 |  |  |  | 1808 |
| 2013 | 7.h | 955 |  | 11355 |  |  | 1728 | 14038 |
| 2013 | 8.a | 1354 |  | 870 |  |  |  | 2224 |
| 2013 | 8.d |  |  | 270 |  |  |  | 270 |
| 2013 | 7.j | 10873 |  | 27003 |  |  | 1653 | 39529 |
| 2014 | ALL | 8758 |  | 34622 |  |  | 38 | 43418 |
| 2014 | 6.a |  |  | 182 |  |  | 30 | 212 |
| 2014 | 7.b | 12 |  | 3262 |  |  |  | 3274 |
| 2014 | 7.g |  |  | 135 |  |  |  | 135 |
| 2014 | 7.h | 4808 |  | 18389 |  |  |  | 23196 |
| 2014 | 8.a |  |  | 119 |  |  |  | 119 |
| 2014 | 7.j | 3886 |  | 12536 |  |  | 8 | 16429 |
| 2014 | 7.k | 53 |  |  |  |  |  | 53 |
| 2015 | ALL | 29 | 5 | 16325 | 375 | 104 |  | 16837 |
| 2015 | $6 . a$ | 10 |  | 116 |  | 9 |  | 134 |
| 2015 | 7.b | 8 | 4 | 2609 |  | 85 |  | 2706 |
| 2015 | 7.c |  |  | 220 |  |  |  | 220 |


| Year | Area | Denmark | Germany | Ireland | The.Netherlands | UKE | UKS | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 7.g |  |  | 547 |  |  |  | 547 |
| 2015 | 7.h | 5 |  | 8506 |  |  |  | 8510 |
| 2015 | 8.a | 6 | 1 | 682 |  |  |  | 688 |
| 2015 | 7.j |  |  | 3646 |  | 10 |  | 3655 |
| 2015 | 6 |  |  |  | 128 |  |  | 128 |
| 2015 | 7 |  |  |  | 33 |  |  | 33 |
| 2015 | 8 |  |  |  | 214 |  |  | 214 |
| 2016 | ALL | 337 | 7 | 17496 | 171 | 21 |  | 18031 |
| 2016 | 6.a |  |  | 377 | 45 |  |  | 422 |
| 2016 | 7.b |  | 5 | 1198 | 35 | 0.66 |  | 1239 |
| 2016 | 7.c |  |  |  | 0.08 |  |  | 0.08 |
| 2016 | 7.e |  |  |  | 0.02 |  |  | 0.02 |
| 2016 | 7.h | 330 |  | 6771 |  |  |  | 7101 |
| 2016 | 7.j |  |  | 1852 | 90 | 16 |  | 1959 |
| 2016 | 8.a | 2 | 1 | 6173 |  | 5 |  | 6181 |
| 2016 | 8.b |  |  |  |  | 0.11 |  | 0.11 |
| 2016 | 8.d | 5 |  | 1124 |  |  |  | 1129 |
| 2017 | ALL | 548 |  | 15485 | 182 | 0.13 |  | 16215 |
| 2017 | 4.a |  |  |  | 0.03 |  |  | 0.03 |
| 2017 | 6.a | 37 |  | 907 | 34 |  |  | 979 |
| 2017 | 7.b |  |  | 124 | 118 |  |  | 242 |
| 2017 | 7.c |  |  |  | 20 |  |  | 20 |
| 2017 | 7.d | 1 |  |  |  |  |  | 1 |
| 2017 | 7.e |  |  |  | 0.08 |  |  | 0.08 |
| 2017 | 7.f |  |  |  |  | 0.02 |  | 0.02 |
| 2017 | 7.g |  |  | 1 |  | 0.02 |  | 1 |
| 2017 | 7.h | 239 |  | 2961 |  | 0.09 |  | 3200 |
| 2017 | 7.j |  |  | 33 | 9 |  |  | 43 |
| 2017 | 8.a | 271 |  | 10543 |  |  |  | 10814 |
| 2017 | 8.d |  |  | 915 |  |  |  | 915 |
| ALL | ALL | 90344 | 12 | 413378 | 727 | 126 | 22128 | 526711 |

Table 3.1.2.3. Boarfish in ICES Subareas 27.6, 7, 8. Discards of boarfish in demersal and non-target pelagic fisheries by year ( $t$ ), 2003-2017. (Data provided by Working Group members). These figures may not in all cases correspond to the official statistics and cannot be used for management purposes.

| Year | Germany | Ireland | Netherlands | Spain | UK | Danemark | Lituania | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 |  | 119 | 1998 | 8812 |  |  |  | 10929 |
| 2004 |  | 60 | 837 | 3579 |  |  |  | 4476 |
| 2005 |  | 55 | 733 | 5007 |  |  |  | 5795 |
| 2006 |  | 22 | 411 | 3933 |  |  |  | 4366 |
| 2007 |  | 549 | 23 | 2617 |  |  |  | 3189 |
| 2008 |  | 920 | 738 | 8410 |  |  |  | 10068 |
| 2009 |  | 377 | 1258 | 5047 |  |  |  | 6682 |
| 2010 |  | 85 | 512 | 5947 |  |  |  | 6544 |
| 2011 | 49 | 107 | 185 | 5461 |  |  |  | 5802 |
| 2012 |  | 181 | 88 | 6365 |  |  |  | 6634 |
| 2013 | 22 | 47 | 11 | 5518 |  |  |  | 5598 |
| 2014 | 117 | 50 | 477 | 1119 | 50 |  |  | 1813 |
| 2015 |  | 7 |  | 921 | 1 |  |  | 929 |
| 2016 | 869 | 20 | 41 | 348 | 4 |  | 1 | 1284 |
| 2017 |  | 640 | 146 |  |  | 386 | 1 | 1173 |

Table 3.2.1.1. Boarfish in ICES Subareas 27.6, 7, 8. General boarfish age length key produced from 2012 commercial samples. Figures highlighted in grey are estimated.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 8 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 8 |  | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 8 |  | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 9 |  | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 |  |  | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 |  |  | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 |  |  | 2 | 10 | 3 |  |  |  |  |  |  |  |  |  |  |
| 11 |  |  | 1 | 29 | 14 | 2 | 2 |  |  |  |  |  |  |  |  |
| 12 |  |  |  | 9 | 21 | 21 | 18 | 2 | 2 | 1 |  |  |  |  |  |
| 12 |  |  |  | 4 | 17 | 22 | 38 | 12 | 8 |  |  |  |  |  | 1 |
| 12 |  |  |  |  | 5 | 9 | 42 | 37 | 14 | 6 | 2 |  | 1 | 1 | 1 |
| 13 |  |  |  |  | 2 | 4 | 31 | 28 | 24 | 12 | 6 | 2 | 3 | 1 | 5 |
| 14 |  |  |  |  | 1 | 3 | 25 | 22 | 21 | 14 | 6 | 5 | 4 | 2 | 11 |
| 14 |  |  |  |  |  |  | 6 | 8 | 18 | 22 | 8 | 3 | 7 | 1 | 20 |
| 14 |  |  |  |  |  | 1 | 1 | 2 | 3 | 8 | 1 | 6 | 6 | 6 | 30 |
| 15 |  |  |  |  |  |  | 1 | 1 |  | 2 | 2 | 2 | 5 | 2 | 19 |
| 16 |  |  |  |  |  |  |  |  |  | 2 |  |  |  | 2 | 19 |
| 16 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 8 |
| 16 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 17 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 18 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 18 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 18 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |

Table 3.2.1.2. Boarfish in ICES Subareas 27.6, 7, 8. Number of samples collected from the catch per year.
$\left.\begin{array}{llclccc}\hline \text { YEAR } & \text { LANDINGS } & \begin{array}{c}\text { \% LANDINGS COVERED BY SAMPLING } \\ \text { PROGRAMME }\end{array} & \begin{array}{c}\text { No. } \\ \text { SAMPLES }\end{array} & \text { No. MEASURED }\end{array} \begin{array}{c}\text { No. } \\ \text { AGED }\end{array}\right]$

Table 3.2.1.3. Boarfish in ICES Subareas 5, 27.6, 7, 8. The allocation of Age length keys to unsampled metiers in 2017.

| Country | Area | QUARTER | LANDED | ALK |
| :---: | :---: | :---: | :---: | :---: |
| DK | 7.d | 1 | 1 | IE_8.d_Q1 IE_8.a_Q1 |
|  |  |  |  | IE_7.j_Q1 IE_7.h_Q1 |
|  |  |  |  | DK_7.h_Q1 DK_8.a_Q1 |
| DK | 7.h | 1 | 239 | IE_7.h_Q1 DK_7.h_Q1 |
| DK | $8 . a$ | 1 | 271 | IE_8.a_Q1 DK_8.a_Q1 |
| IE | 7.b | 1 | 95 | IE_7.j_Q1 |
| IE | 7.b | 4 | 29 | IE_7.h_Q4 |
| IE | 7.g | 4 | 1 | IE_7.h_Q3 IE_7.h_Q4 |
| IE | 7.h | 1 | 188 | IE_7.h_Q1 DK_7.h_Q1 |
| IE | 7.h | 3 | 95 | IE_7.h_Q3 |
| IE | 7.h | 4 | 2678 | IE_7.h_Q4 |
| IE | 7.j | 1 | 33 | IE_7.j_Q1 |
| IE | 8.a | 1 | 7357 | IE_8.a_Q1 DK_8.a_Q1 |
| IE | 8.a | 3 | 50 | IE_8.a_Q3 |
| IE | 8.a | 4 | 3135 | IE_8.a_Q4 |
| IE | 8.d | 1 | 915 | IE_8.d_Q1 |
| NL | 7.b | 1 | 65 | IE_7.j_Q1 |
| NL | 7.b | 2 | 0.42 | IE_7.j_Q1 |
| NL | 7.b | 3 | 53 | IE_7.j_Q1 |
| NL | 7.c | 4 | 20 | IE_7.h_Q4 |
| NL | 7.e | 1 | 0.08 | IE_8.a_Q1 IE_7.h_Q1 <br> DK_7.h_Q1 DK_8.a_Q1 |
| NL | 7.j | 1 | 0.01 | IE_7.j_Q1 |
| NL | 7.j | 2 | 1 | IE_7.j_Q1 |
| NL | 7.j | 3 | 8 | IE_7.h_Q3 IE_7.h_Q4 |
| UKE | $7 . f$ | 2 | 0.02 | IE_7.j_Q1 IE_7.h_Q1 <br> IE_7.h_Q3 DK_7.h_Q1 |
| UKE | 7.g | 2 | 0.02 | IE_7.j_Q1 IE_7.h_Q1 <br> IE_7.h_Q3 DK_7.h_Q1 |
| UKE | 7.h | 2 | 0.09 | IE_7.h_Q1 IE_7.h_Q3 DK_7.h_Q1 |

Table 3.2.1.4. Boarfish in ICES Subareas 27.6, 7, 8. Catch per country and corresponding number of samples collected in 2017.

| Country | OfFICIAL.CATCH | \%.LANDINGS.covered | No.SAMPLEs | No.MEASURED | No.AGED |
| :--- | :--- | :--- | :--- | :--- | :--- |
| DK | 548 | 4 | 374 |  |  |
| ES | 640 |  |  |  |  |
| IE | 15631 |  |  | 766 |  |
| NL | 182 |  |  |  |  |
| UKE | 386 |  |  |  |  |
| UKS | 1 |  |  |  |  |
| Total |  |  |  |  |  |

Table 3.2.1.5. Boarfish in ICES Subareas 27.6, 7, 8. Catch per area and corresponding number of samples collected in 2017.

| Area | Official.catch | No.samples | No.measured | No.measured.per.1000t |
| :---: | :---: | :---: | :---: | :---: |
| 27.4.a | 0.03 |  |  |  |
| 27.6.a | 980 |  |  |  |
| 27.6.b | 5 |  |  |  |
| 27.7.b | 276 |  |  |  |
| 27.7.c | 81 |  |  |  |
| 27.7.d | 1 |  |  |  |
| 27.7.e | 371 |  |  |  |
| 27.7.f | 2 |  |  |  |
| 27.7.g | 4 |  |  |  |
| 27.7.h | 3363 | 7 | 452 | 134 |
| 27.8.a | 10814 | 9 | 595 | 55 |
| 27.8.b | 6 |  |  |  |
| 27.8.c | 208 |  |  |  |
| 27.8.d | 915 | 1 | 24 | 26 |
| 27.7.j | 361 | 1 | 69 | 191 |
| 27.7.k | 1 |  |  |  |

Table 3.2.1.6. Boarfish in ICES Subareas 27.6, 7, 8. Proxy catch numbers-at-age of the international catches (raised numbers in ‘000s) for the years 2007-2017

|  | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  |  | 1575 | 2415 |  | 28 | 301 |  | 5556 | 218 | 1862 |
| 2 | 352 | 5488 | 15043 | 11229 | 2894 | 893 | 7148 | 695 | 116135 | 2385 | 4387 |
| 3 | 2114 | 21140 | 65744 | 72709 | 41913 | 5467 | 156680 | 49503 | 32248 | 10737 | 8830 |
| 4 | 40851 | 105575 | 338931 | 294382 | 28148 | 41278 | 58522 | 127520 | 16588 | 25114 | 34448 |
| 5 | 48915 | 141300 | 475619 | 567689 | 30116 | 110272 | 59797 | 93705 | 24564 | 20263 | 27266 |
| 6 | 62713 | 195339 | 543707 | 878363 | 175696 | 146582 | 68949 | 67275 | 26566 | 18025 | 21103 |
| 7 | 26132 | 104031 | 307333 | 522703 | 143967 | 492078 | 302967 | 193061 | 74115 | 61229 | 55189 |
| 8 | 29766 | 66570 | 172783 | 293719 | 107126 | 365840 | 250341 | 139124 | 52052 | 47573 | 38229 |
| 9 | 56075 | 53159 | 155477 | 276672 | 77861 | 271916 | 212318 | 121042 | 44615 | 42478 | 32258 |
| 10 | 44875 | 46893 | 130148 | 232122 | 60022 | 173486 | 160137 | 94225 | 34264 | 35150 | 25716 |
| 11 | 14019 | 15289 | 42521 | 78588 | 46079 | 69396 | 63025 | 36078 | 12999 | 13297 | 9560 |
| 12 | 32359 | 21178 | 61350 | 114600 | 40468 | 40968 | 41490 | 24895 | 9114 | 9132 | 7564 |
| 13 | 4848 | 11854 | 39609 | 59932 | 24352 | 58888 | 59380 | 36309 | 13362 | 13774 | 10922 |
| 14 | 16837 | 13570 | 31569 | 59060 | 19724 | 30277 | 30355 | 19064 | 7152 | 6682 | 5924 |
| 15+ | 109481 | 112947 | 196967 | 349320 | 157707 | 217260 | 239366 | 150688 | 59139 | 49589 | 40797 |

Table 3.2.2.1. Boarfish in ICES Subareas 27.6, 7, 8. Length-frequency distributions of the international catches (raised numbers in '000s) for the years 2007-2017.

| TL (см) | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4.5 |  |  |  |  |  |  |  |  | 14 |  |  | 14 |
| 5 |  |  |  |  |  |  |  |  | 878 |  |  | 878 |
| 5.5 |  |  |  |  |  |  |  |  | 515 |  |  | 515 |
| 6 |  |  |  | 156 |  |  |  |  | 810 |  | 765 | 1731 |
| 6.5 |  |  |  | 439 |  |  |  |  | 14 |  | 4607 | 5060 |
| 7 |  |  |  | 1090 | 522 | 56 | 52 |  | 513 | 417 | 5250 | 7900 |
| 7.5 |  |  | 1354 | 1574 |  |  | 551 |  | 10598 | 1684 | 12616 | 28377 |
| 8 |  |  | 677 | 375 | 1345 | 185 | 1419 |  | 80716 | 8685 | 11473 | 104875 |
| 8.5 |  |  |  | 1082 |  | 555 | 3592 | 1064 | 49508 | 6412 | 10115 | 72328 |
| 9 |  |  | 677 | 5382 | 851 | 555 | 7263 | 327 | 10219 | 7104 | 3874 | 36252 |
| 9.5 |  | 7473 | 17367 | 7883 | 7012 | 641 | 47509 | 4916 | 213 | 23065 | 14047 | 130126 |
| 10 | 9609 | 11209 | 54130 | 29410 | 33243 | 2791 | 94702 | 31649 | 1211 | 46010 | 32346 | 346310 |
| 10.5 |  | 52308 | 174796 | 130889 | 15848 | 6132 | 59833 | 71344 | 3865 | 39071 | 36242 | 590328 |
| 11 | 84555 | 63517 | 343283 | 361774 | 70615 | 24571 | 18359 | 108261 | 12226 | 14181 | 32445 | 1133787 |
| 11.5 |  | 59781 | 321637 | 655875 | 93487 | 81928 | 20938 | 82470 | 28142 | 18249 | 31589 | 1394096 |
| 12 | 44199 | 119561 | 297737 | 739025 | 189434 | 264888 | 98564 | 84288 | 41613 | 30975 | 33618 | 1943902 |
| 12.5 |  | 70990 | 207739 | 564347 | 114904 | 398772 | 204868 | 112826 | 42461 | 51110 | 41650 | 1809667 |
| 13 | 82633 | 52308 | 147965 | 353484 | 133539 | 419060 | 315063 | 172416 | 59990 | 57000 | 46495 | 1839953 |
| 13.5 |  | 29890 | 149314 | 246146 | 51235 | 307533 | 285688 | 153742 | 52625 | 58696 | 43121 | 1377990 |
| 14 | 117224 | 22418 | 105782 | 224611 | 50857 | 176710 | 210137 | 138549 | 50139 | 76872 | 45353 | 1218652 |
| 14.5 |  | 14945 | 71273 | 127711 | 25309 | 89726 | 105571 | 74059 | 28771 | 37755 | 39524 | 614644 |
| 15 | 65338 | 33627 | 47816 | 125463 | 25569 | 52791 | 62175 | 43347 | 16087 | 23137 | 21854 | 517204 |
| 15.5 |  | 11209 | 13082 | 81386 | 5473 | 25065 | 31122 | 22629 | 8572 | 7841 | 4932 | 211311 |
| 16 | 13452 | 11209 | 19397 | 24256 | 4181 | 13149 | 14990 | 7672 | 4331 | 625 | 1020 | 114282 |
| 16.5 |  | 3736 | 4061 | 6209 | 2280 | 2738 | 4918 | 2134 | 2081 | 128 |  | 28285 |
| 17 |  | 3736 | 677 | 1913 | 456 | 827 | 1109 | 1361 | 289 |  |  | 10368 |
| 17.5 |  |  |  |  |  |  | 407 |  | 23 |  |  | 430 |
| 18 |  |  |  | 283 |  |  | 296 |  |  |  |  | 579 |


| TL (см) | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18.5 |  |  |  |  |  |  | 592 |  |  |  |

Table 3.3.1.1. Boarfish in ICES Subareas 27.6. 7, 8. Acoustic survey abundance and biomass estimates from 2011-2018

|  | Abundance |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age.(Yrs) | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| 0 | - | - | - | - | - | - | - | - |
| 1 | 5 | 21.5 | - | - | 198.5 | 4.6 | 110.9 | 76.7 |
| 2 | 11.6 | 10.8 | 78 | - | 319.2 | 35.7 | 126.7 | 31.2 |
| 3 | 57.8 | 174.1 | 1842.9 | 15 | 16.6 | 45.5 | 344.6 | 115 |
| 4 | 187.4 | 64.8 | 696.4 | 98.2 | 34.3 | 43.6 | 367.3 | 68.3 |
| 5 | 436.7 | 95 | 381.6 | 102.3 | 80 | 6 | 156 | 106.7 |
| 6 | 1165.9 | 736.1 | 253.8 | 104.9 | 112 | 10 | 209 | 165.9 |
| 7 | 1184.2 | 973.8 | 1056.6 | 414.6 | 437.4 | 169 | 493.1 | 320.7 |
| 8 | 703.6 | 758.9 | 879.4 | 343.8 | 362.9 | 112.6 | 468.3 | 197.7 |
| 9 | 1094.5 | 848.6 | 800.9 | 341.9 | 353.5 | 117.6 | 397.2 | 293.4 |
| 10 | 1031.5 | 955.9 | 703.8 | 332.3 | 360 | 96.6 | 285.8 | 624.7 |
| 11 | 332.9 | 650.9 | 263.7 | 129.9 | 131.7 | 17 | 120.9 | 339.2 |
| 12 | 653.3 | 1099.7 | 202.9 | 104.9 | 113 | 32 | 82.1 | 264.1 |
| 13 | 336 | 857.2 | 296.6 | 166.4 | 174 | 48.7 | 74.4 | 198.4 |
| 14 | 385 | 655.8 | 169.8 | 88.5 | 108 | 18.3 | 220.4 | 116.5 |
| 15+ | 3519 | 6353.7 | 1464.3 | 855.1 | 1195 | 400.1 | 931 | 302.4 |
| $\begin{gathered} \hline \text { TSN } \\ (\prime 000) \end{gathered}$ | 11104 | 14257 | 9091 | 3098 | 3996 | 1157 | 4387 | 3221 |
| TSB (t) | 670176 | 863446 | 439890 | 187779 | 232634 | 69690 | 230062 | 186252 |
| SSB (t) | 669392 | 861544 | 423158 | 187654 | 226659 | 69103 |  | 184624 |
| CV | 21.2 | 10.6 | 17.5 | 15.1 | 17 | 16.4 | 21.9 | 19.9 |

Table 3.3.2.1. Boarfish in ICES Subareas 27.6, 7, 8. IBTS length-frequency data

| SURVE Y | $\begin{gathered} Y_{\text {EA }} \\ \mathrm{R} \end{gathered}$ | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 2 0 | $\begin{gathered} \mathrm{M} \\ \mathrm{~L} \end{gathered}$ | $\begin{gathered} \text { ML.MAT } \\ \text { URE } \\ \hline \end{gathered}$ | TOT AL | TOTAL.MAT <br> URE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ЕVHO <br> E | $\begin{aligned} & 199 \\ & 7 \end{aligned}$ |  | 5 | 11 | 7 | 17 | 197 | 2659 | 5020 | 3719 | 3598 | 4429 | $\begin{aligned} & 1206 \\ & 5 \end{aligned}$ | $\begin{aligned} & 1665 \\ & 1 \end{aligned}$ | 7198 | 3455 | 501 | 18 | 1 |  |  | 12 | 13 | $\begin{aligned} & 5954 \\ & 8 \end{aligned}$ | 47915 |
| EVHO <br> E | $\begin{aligned} & 199 \\ & 8 \end{aligned}$ |  | 1 | 4 | 26 | 76 | 2093 | $\begin{aligned} & 1828 \\ & 3 \end{aligned}$ | 8631 | 6125 | 5966 | 7095 | $\begin{aligned} & 1173 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1407 \\ & 8 \end{aligned}$ | 9260 | 5076 | 934 | 8 |  |  | 1 | 11 | 13 | $\begin{aligned} & 8938 \\ & 7 \end{aligned}$ | 54148 |
| ЕУно <br> E | $\begin{aligned} & 199 \\ & 9 \end{aligned}$ |  |  | 13 | 52 | 33 | 245 | $\begin{aligned} & 1117 \\ & 7 \end{aligned}$ | $\begin{aligned} & 2661 \\ & 0 \end{aligned}$ | $\begin{aligned} & 2394 \\ & 7 \end{aligned}$ | 6684 | 2899 | 4709 | 7868 | 6160 | 1353 | 267 | 7 |  |  |  | 10 | 12 | $\begin{aligned} & 9202 \\ & 3 \end{aligned}$ | 29947 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 200 \\ & 0 \end{aligned}$ |  | 17 | 79 | 120 | 8 | 1504 | $\begin{aligned} & 2689 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1767 \\ & 4 \end{aligned}$ | 9836 | $\begin{aligned} & 2196 \\ & 7 \end{aligned}$ | $\begin{aligned} & 1638 \\ & 2 \end{aligned}$ | $\begin{aligned} & 2958 \\ & 5 \end{aligned}$ | $\begin{aligned} & 3685 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1652 \\ & 2 \end{aligned}$ | 5397 | 989 | 75 |  |  |  | 11 | 12 | $\begin{aligned} & 1839 \\ & 03 \end{aligned}$ | 127769 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 200 \\ & 1 \end{aligned}$ |  | 1 | 45 | 687 | 489 | 913 | $\begin{aligned} & 2129 \\ & 7 \end{aligned}$ | $3717$ | $\begin{aligned} & 1327 \\ & 6 \end{aligned}$ | $\begin{aligned} & 2835 \\ & 5 \end{aligned}$ | $\begin{aligned} & 3151 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1830 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1223 \\ & 2 \end{aligned}$ | 6471 | 3186 | 1270 | 81 | 4 |  |  | 10 | 12 | $\begin{aligned} & 1753 \\ & 03 \end{aligned}$ | 101422 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 200 \\ & 2 \end{aligned}$ |  | 2 | 18 | 23 | 11 | 547 | 9631 | $\begin{aligned} & 2987 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1777 \\ & 7 \end{aligned}$ | $\begin{aligned} & 1329 \\ & 0 \end{aligned}$ | 9470 | 9697 | 9751 | 6268 | 2484 | 641 | 37 | 1 | 1 |  | 10 | 12 | $\begin{aligned} & 1095 \\ & 22 \end{aligned}$ | 51639 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 200 \\ & 3 \end{aligned}$ |  |  | 17 | 47 | 17 | 57 | 426 | 1655 | 7142 | $\begin{aligned} & 2001 \\ & 8 \end{aligned}$ | $\begin{aligned} & 2484 \\ & 2 \end{aligned}$ | $\begin{aligned} & 2098 \\ & 9 \end{aligned}$ | $\begin{aligned} & 2126 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1449 \\ & 4 \end{aligned}$ | 7086 | 1550 | 36 |  |  |  | 12 | 12 | $\begin{aligned} & 1196 \\ & 39 \end{aligned}$ | 110277 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 200 \\ & 4 \end{aligned}$ |  |  | 33 | 512 | 378 | 123 | 1248 | 1419 | 1307 | 1083 | 3102 | 7308 | 7224 | 6353 | 7866 | 3630 | 241 | 5 |  |  | 13 | 14 | $\begin{aligned} & 4183 \\ & 3 \end{aligned}$ | 36813 |
| EVHO <br> E | $\begin{aligned} & 200 \\ & 5 \end{aligned}$ |  | 2 | 93 | 975 | 1285 | 146 | 1100 | 2326 | 1229 | 1553 | 3183 | $\begin{aligned} & 1339 \\ & 8 \end{aligned}$ | $\begin{aligned} & 1575 \\ & 8 \end{aligned}$ | 9834 | 6010 | 1658 | 117 | 70 |  |  | 12 | 13 | $\begin{aligned} & 5873 \\ & 8 \end{aligned}$ | 51580 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 200 \\ & 6 \end{aligned}$ | 1 | 26 | 112 | 79 | 75 | $\begin{aligned} & 1551 \\ & 0 \end{aligned}$ | $\begin{aligned} & 3756 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1075 \\ & 0 \end{aligned}$ | 3622 | 2127 | 1521 | 1955 | 4131 | 3955 | 2535 | 921 | 94 | 2 | 12 |  | 8 | 13 | $\begin{aligned} & 8499 \\ & 4 \end{aligned}$ | 17253 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 200 \\ & 7 \end{aligned}$ |  | 8 | 187 | 467 | 234 | 1503 | $\begin{aligned} & 2268 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1260 \\ & 65 \end{aligned}$ | $\begin{aligned} & 6453 \\ & 6 \end{aligned}$ | 6341 | 6731 | 5431 | 6004 | 5911 | 4238 | 1409 | 118 | 11 |  |  | 9 | 12 | $\begin{aligned} & 2518 \\ & 82 \end{aligned}$ | 36193 |
| $\begin{aligned} & \text { ЕVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 200 \\ & 8 \end{aligned}$ |  | 3 | 434 | 2807 | 827 | 5341 | $\begin{aligned} & 5318 \\ & 9 \end{aligned}$ | $\begin{aligned} & 2472 \\ & 96 \end{aligned}$ | $\begin{aligned} & 1653 \\ & 92 \end{aligned}$ | $\begin{aligned} & 1632 \\ & 00 \end{aligned}$ | $\begin{aligned} & 6938 \\ & 2 \end{aligned}$ | $\begin{aligned} & 3843 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1839 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1725 \\ & 8 \end{aligned}$ | 9178 | 3490 | 745 | 6 | 1 |  | 9 | 11 | $\begin{aligned} & 7953 \\ & 71 \end{aligned}$ | 320083 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 200 \\ & 9 \end{aligned}$ |  | 6 | 128 | 194 | 72 | 1496 | $\begin{aligned} & 1976 \\ & 9 \end{aligned}$ | $\begin{aligned} & 3581 \\ & 9 \end{aligned}$ | 5264 | 3913 | 9556 | $\begin{aligned} & 1226 \\ & 9 \end{aligned}$ | 9402 | $\begin{aligned} & 1083 \\ & 1 \end{aligned}$ | 6720 | 775 | 38 | 1 |  |  | 10 | 13 | $\begin{aligned} & 1162 \\ & 52 \end{aligned}$ | 53505 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 201 \\ & 0 \end{aligned}$ |  | 21 | 529 | 116 | 154 | 5755 | $\begin{aligned} & 4643 \\ & 8 \end{aligned}$ | $\begin{aligned} & 7498 \\ & 6 \end{aligned}$ | $\begin{aligned} & 2717 \\ & 5 \end{aligned}$ | $\begin{aligned} & 1195 \\ & 2 \end{aligned}$ | $\begin{aligned} & 3742 \\ & 0 \end{aligned}$ | $\begin{aligned} & 5831 \\ & 3 \end{aligned}$ | $\begin{aligned} & 3473 \\ & 7 \end{aligned}$ | $\begin{aligned} & 3377 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1462 \\ & 6 \end{aligned}$ | 1561 | 249 | 8 | 1 |  | 10 | 12 | $\begin{aligned} & 3478 \\ & 14 \end{aligned}$ | 192641 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 201 \\ & 1 \end{aligned}$ |  | 60 | 95 | 215 | 5 | 541 | 2247 | 8368 | $\begin{aligned} & 1525 \\ & 6 \end{aligned}$ | $\begin{aligned} & 3322 \\ & 1 \end{aligned}$ | $\begin{aligned} & 3023 \\ & 7 \end{aligned}$ | $\begin{aligned} & 5038 \\ & 4 \end{aligned}$ | $\begin{aligned} & 5655 \\ & 9 \end{aligned}$ | $\begin{aligned} & 3667 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1186 \\ & 7 \end{aligned}$ | 3082 | 573 | $\begin{aligned} & 15 \\ & 9 \end{aligned}$ | 47 |  | 12 | 12 | $\begin{aligned} & 2495 \\ & 90 \end{aligned}$ | 222803 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 201 \\ & 2 \end{aligned}$ |  | 9 | 145 | 584 | 137 | 2922 | $\begin{aligned} & 2886 \\ & 5 \end{aligned}$ | $\begin{aligned} & 2681 \\ & 6 \end{aligned}$ | 6124 | $\begin{aligned} & 1173 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1360 \\ & 6 \end{aligned}$ | $\begin{aligned} & 2236 \\ & 9 \end{aligned}$ | $\begin{aligned} & 3713 \\ & 5 \end{aligned}$ | $\begin{aligned} & 4408 \\ & 2 \end{aligned}$ | $\begin{aligned} & 1996 \\ & 3 \end{aligned}$ | 4893 | 127 | 1 |  |  | 11 | 13 | $\begin{aligned} & 2195 \\ & 16 \end{aligned}$ | 153914 |
| ЕУно <br> E | $\begin{aligned} & 201 \\ & 3 \end{aligned}$ |  | 3 | 48 | 91 | 10 | 306 | 2185 | 2165 | 2542 | $\begin{aligned} & 1364 \\ & 9 \end{aligned}$ | 9932 | $\begin{aligned} & 1498 \\ & 7 \end{aligned}$ | $\begin{aligned} & 3775 \\ & 5 \end{aligned}$ | $\begin{aligned} & 4052 \\ & 4 \end{aligned}$ | $\begin{aligned} & 2010 \\ & 7 \end{aligned}$ | 6918 | 666 |  | 2 |  | 13 | 13 | $\begin{aligned} & 1518 \\ & 90 \end{aligned}$ | 144540 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 201 \\ & 4 \end{aligned}$ |  | 2 | 693 | 1386 | 508 | 84 | 1440 | 885 | 3074 | 8732 | $\begin{aligned} & 2858 \\ & 6 \end{aligned}$ | $\begin{aligned} & 3939 \\ & 7 \end{aligned}$ | $\begin{aligned} & 7412 \\ & 2 \end{aligned}$ | $\begin{aligned} & 6973 \\ & 6 \end{aligned}$ | $\begin{aligned} & 2687 \\ & 1 \end{aligned}$ | 3908 | 59 | $\begin{aligned} & 43 \\ & 3 \end{aligned}$ |  |  | 13 | 13 | $\begin{aligned} & 2599 \\ & 15 \end{aligned}$ | 251844 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 201 \\ & 5 \end{aligned}$ |  | 5 | 183 | 5898 | 4143 | 607 | $\begin{aligned} & 1907 \\ & 5 \end{aligned}$ | $\begin{aligned} & 1792 \\ & 69 \end{aligned}$ | $\begin{aligned} & 1190 \\ & 04 \end{aligned}$ | $\begin{aligned} & 1576 \\ & 5 \end{aligned}$ | $\begin{aligned} & 1801 \\ & 4 \end{aligned}$ |  | $\begin{aligned} & 6202 \\ & 4 \end{aligned}$ | $\begin{aligned} & 5990 \\ & 4 \end{aligned}$ | $\begin{aligned} & 2152 \\ & 5 \end{aligned}$ | 5487 | 541 | $\begin{aligned} & 42 \\ & 9 \end{aligned}$ | 8 |  | 10 | 13 | $\begin{aligned} & 5734 \\ & 55 \end{aligned}$ | 245271 |
| $\begin{aligned} & \text { EVHO } \\ & \text { E } \end{aligned}$ | $\begin{aligned} & 201 \\ & 6 \end{aligned}$ | 5 | 31 | 379 | 846 | 115 | 733 | $\begin{aligned} & 1028 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1428 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1725 \\ & 1 \end{aligned}$ | $\begin{aligned} & 4213 \\ & 2 \end{aligned}$ | $\begin{aligned} & 2530 \\ & 4 \end{aligned}$ | $\begin{aligned} & 6858 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1306 \\ & 33 \end{aligned}$ | $\begin{aligned} & 1312 \\ & 20 \end{aligned}$ | $\begin{aligned} & 4853 \\ & 8 \end{aligned}$ | $\begin{aligned} & 1161 \\ & 1 \end{aligned}$ | $\begin{aligned} & 135 \\ & 8 \end{aligned}$ | 26 |  |  | 13 | 13 | $\begin{aligned} & 5033 \\ & 29 \end{aligned}$ | 459405 |
| EVHO | 201 |  | 2 | 103 | 129 | 3 | 27 | 269 | 198 | 5 |  |  |  |  |  |  |  |  |  |  |  | 6 |  | 735 |  |


| SURVE Y | $\begin{gathered} \text { YEA } \\ \mathrm{R} \end{gathered}$ | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | $\begin{aligned} & 2 \\ & 0 \end{aligned}$ | м | $\begin{aligned} & \text { ML.MAT } \\ & \text { URE } \end{aligned}$ | $\begin{aligned} & \text { TOT } \\ & \text { AL } \end{aligned}$ | TOTAL.MAT URE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| E | 7 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| IGFS | $\begin{aligned} & 200 \\ & 3 \end{aligned}$ |  | 1 | 32 | 22 | 7 | 22 | 129 | 172 | 879 | 2942 | 2322 | 1326 | 3822 | 4628 | 2898 | 896 | 163 | 38 |  |  | 13 | 13 | $\begin{aligned} & 2029 \\ & 9 \end{aligned}$ | 19035 |
| IGFS | $\begin{aligned} & 200 \\ & 4 \end{aligned}$ |  | 23 | 63 | 34 | 8 | 96 | 532 | 1431 | 369 | 344 | 410 | 2253 | 4320 | 4698 | 3966 | 1017 | 87 | 2 | 1 |  | 13 | 14 | $\begin{aligned} & 1965 \\ & 4 \end{aligned}$ | 17098 |
| IGFS | $\begin{aligned} & 200 \\ & 5 \end{aligned}$ |  | 8 | 59 | 52 | 20 | 203 | 1024 | 585 | 288 | 636 | 341 | 3463 | $\begin{aligned} & 1145 \\ & 7 \end{aligned}$ | $\begin{aligned} & 1134 \\ & 8 \end{aligned}$ | 7955 | 1744 | 382 | 2 | $\begin{aligned} & 0.9 \\ & 7 \end{aligned}$ |  | 13 | 14 | $\begin{aligned} & 3956 \\ & 9 \end{aligned}$ | 37330 |
| IGFS | $\begin{aligned} & 200 \\ & 6 \end{aligned}$ | 5 | 60 | 68 | 48 | 35 | 212 | 969 | 621 | 2046 | 4190 | 8044 | 7946 | $\begin{aligned} & 2420 \\ & 8 \end{aligned}$ | $\begin{aligned} & 4211 \\ & 9 \end{aligned}$ | $\begin{aligned} & 3216 \\ & 8 \end{aligned}$ | $\begin{aligned} & 1229 \\ & 6 \end{aligned}$ | ${ }_{4}^{245}$ | $\begin{aligned} & 53 \\ & 2 \end{aligned}$ |  |  | 14 | 14 | $\begin{aligned} & 1380 \\ & 21 \end{aligned}$ | 133957 |
| IGFS | $\begin{aligned} & 200 \\ & 7 \end{aligned}$ | 1 | 6 | 44 | 18 | 31 | 501 | 923 | 1251 | 1638 | 1166 | 2510 | 3581 | 8275 | $\begin{aligned} & 1074 \\ & 0 \end{aligned}$ | 7093 | 1934 | 92 |  |  |  | 13 | 14 | $\begin{aligned} & 3980 \\ & 4 \end{aligned}$ | 35391 |
| IGFS | $\begin{aligned} & 200 \\ & 8 \end{aligned}$ |  |  | 26 | 18 | 23 | 127 | 672 | 531 | 2095 | $\begin{aligned} & 1378 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1766 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1926 \\ & 8 \end{aligned}$ | $\begin{aligned} & 1698 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1948 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1595 \\ & 3 \end{aligned}$ | 8789 | $\begin{aligned} & 174 \\ & 7 \end{aligned}$ | 76 | 1 |  | 13 | 13 | $\begin{aligned} & 1172 \\ & 31 \end{aligned}$ | 113741 |
| IGFS | $\begin{aligned} & 200 \\ & 9 \end{aligned}$ |  | 3 | 80 | 76 | 25 | 94 | 228 | 486 | 1000 | 1139 | 9081 | 7749 | 5138 | 6921 | 5592 | 1084 | 68 | 1 |  |  | 12 | 13 | $\begin{aligned} & 3876 \\ & 3 \end{aligned}$ | 36772 |
| IGFS | $\begin{aligned} & 201 \\ & 0 \end{aligned}$ |  | 6 | 42 | 3 | 18 | 199 | 272 | 463 | 920 | 393 | 7914 | $\begin{aligned} & 3423 \\ & 6 \end{aligned}$ | $2861$ | $\begin{aligned} & 1606 \\ & 3 \end{aligned}$ | 8161 | 1974 | 433 |  |  |  | 13 | 13 | $9970$ | 97784 |
| IGFS | $\begin{aligned} & 201 \\ & 1 \end{aligned}$ |  | 6 | 14 | 5 | 4 | 189 | 772 | 586 | 555 | 670 | 2578 | $\begin{aligned} & 2017 \\ & 1 \end{aligned}$ | $\begin{aligned} & 2208 \\ & 2 \end{aligned}$ | $\begin{aligned} & 1082 \\ & 9 \end{aligned}$ | 5298 | 2207 | 266 | 9 | 6 |  | 13 | 13 | $\begin{aligned} & 6624 \\ & 7 \end{aligned}$ | 64116 |
| IGFS | $\begin{aligned} & 201 \\ & 2 \end{aligned}$ |  | 7 | 36 | 20 | 10 | 131 | 271 | 378 | 702 | 2144 | 1183 | $\begin{aligned} & 1110 \\ & 5 \end{aligned}$ | $\begin{aligned} & 3401 \\ & 0 \end{aligned}$ | $\begin{aligned} & 2274 \\ & 2 \end{aligned}$ | $\begin{aligned} & 1090 \\ & 6 \end{aligned}$ | 3903 | 525 | 4 |  |  | 13 | 13 | $\begin{aligned} & 8807 \\ & 7 \end{aligned}$ | 86521 |
| IGFS | $\begin{aligned} & 201 \\ & 3 \end{aligned}$ | 1 | 3 | 9 | 9 | 20 | 127 | 352 | 340 | 1320 | 2833 | 3971 | $\begin{aligned} & 1557 \\ & 2 \end{aligned}$ | $5163$ | $\begin{aligned} & 5286 \\ & 8 \end{aligned}$ | $\begin{aligned} & 2048 \\ & 5 \end{aligned}$ | 6560 | 492 | 20 |  |  | 14 | 14 | $\begin{aligned} & 1566 \\ & 20 \end{aligned}$ | 154439 |
| IGFS | $\begin{aligned} & 201 \\ & 4 \end{aligned}$ |  | 10 | 68 | 54 | 4 | 18 | 13 | 25 | 60 | 130 | 1127 | 3251 | $\begin{aligned} & 1912 \\ & 5 \end{aligned}$ | $\begin{aligned} & 2301 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1035 \\ & 5 \end{aligned}$ | 2988 | 284 | 18 |  |  | 14 | 14 | $\begin{aligned} & 6054 \\ & 7 \end{aligned}$ | 60295 |
| IGFS | $\begin{aligned} & 201 \\ & 5 \end{aligned}$ |  | 3 | 11 | 16 | 24 | 193 | 1008 | 3708 | 848 | 105 | 713 | 6314 | $\begin{aligned} & 2972 \\ & 7 \end{aligned}$ | $\begin{aligned} & 4822 \\ & 1 \end{aligned}$ | $\begin{aligned} & 3302 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1735 \\ & 0 \end{aligned}$ | $\begin{aligned} & 188 \\ & 5 \end{aligned}$ | $\begin{aligned} & 53 \\ & 1 \end{aligned}$ |  |  | 14 | 14 | $\begin{aligned} & 1436 \\ & 81 \end{aligned}$ | 137870 |
| IGFS | $\begin{aligned} & 201 \\ & 6 \end{aligned}$ | 4 | 31 | 121 | 63 | 7 | 67 | 186 | 1515 | 4057 | 2891 | 1349 | 4110 | $\begin{aligned} & 3275 \\ & 3 \end{aligned}$ | $\begin{aligned} & 5775 \\ & 3 \end{aligned}$ | $\begin{aligned} & 4090 \\ & 7 \end{aligned}$ | $\begin{aligned} & 1552 \\ & 7 \end{aligned}$ | $\begin{aligned} & 367 \\ & 0 \end{aligned}$ | 86 |  |  | 14 | 14 | $\begin{aligned} & 1650 \\ & 97 \end{aligned}$ | 159046 |
| IGFS | $\begin{aligned} & 201 \\ & 7 \end{aligned}$ |  | 6 | 53 | $\begin{aligned} & 1016 \\ & 9 \end{aligned}$ | $\begin{aligned} & 6899 \\ & 15 \end{aligned}$ | 6406 | 1751 | 715 | $\begin{aligned} & 1181 \\ & 8 \end{aligned}$ | $\begin{aligned} & 2188 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1016 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1184 \\ & 1 \end{aligned}$ | $\begin{aligned} & 2558 \\ & 8 \end{aligned}$ | ${ }_{1}^{4231}$ | $\begin{aligned} & 3504 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1711 \\ & 0 \end{aligned}$ | $\begin{aligned} & 329 \\ & 9 \end{aligned}$ | $\begin{aligned} & 36 \\ & 9 \end{aligned}$ |  |  | 7 | 14 | $\begin{aligned} & 8884 \\ & 49 \end{aligned}$ | 167616 |
| $\begin{aligned} & \text { SPNG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 1 \end{aligned}$ |  | 1 |  |  | 31 | 690 | 1311 | 313 | 49 | 9 | 6 | 7 | 7 | 4 |  |  |  | 6 |  |  | 7 | 13 | 2433 | 39 |
| $\begin{aligned} & \text { SPNG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 2 \end{aligned}$ |  | 57 | 38 | 9 | 178 | 3290 | 2743 | 282 | 48 | 10 | 8 | 69 | 162 | 390 | 779 | 246 | 95 |  |  |  | 8 | 15 | 8404 | 1760 |
| $\begin{aligned} & \text { SPNG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 3 \end{aligned}$ |  | 57 | $\begin{aligned} & 120 \\ & 6 \end{aligned}$ | 488 | 97 | 3730 | 3753 | 421 | 105 | 54 | 7 | 4 | 8 | 3 | 2 |  |  |  |  |  | 6 | 11 | 9934 | 77 |
| $\begin{aligned} & \text { SPNG } \\ & \text { FS } \end{aligned}$ | $199$ | 1 | 40 | 33 |  | 342 | 4789 | $\begin{aligned} & 1016 \\ & 2 \end{aligned}$ | 8920 | 3195 | 53 | 106 | 20 | 9 | 12 | 1 |  |  |  |  |  | 7 | 11 | $2768$ | 202 |
| $\begin{aligned} & \text { SPNG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 5 \end{aligned}$ |  | 84 | 108 | 4 | 342 | 3063 | 2157 | 220 | 84 | 65 | 58 | 105 | 105 | 90 | 20 | 4 |  |  |  |  | 7 | 12 | 6510 | 447 |
| $\begin{aligned} & \text { SPNG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 21 \\ & 8 \end{aligned}$ | 537 | 143 | 245 | 4457 | 4449 | 267 | 820 | 722 | 82 | 145 | 126 | 219 | 96 | 39 | 2 |  |  |  | 7 | 12 | $\begin{aligned} & 1256 \\ & 6 \end{aligned}$ | 1431 |
| SPNG | 199 | 2 | 10 | 809 | 441 | 235 | 3458 | 6824 | 2189 | 1923 | 534 | 156 | 353 | 161 | 88 | 3 |  |  |  |  |  | 7 | 11 | 1727 | 1295 |


| Surve Y | $\begin{gathered} \text { Yea } \\ R \end{gathered}$ | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | $\begin{aligned} & 2 \\ & 0 \end{aligned}$ | $\begin{gathered} \mathrm{M} \\ \mathrm{~L} \end{gathered}$ | ML.MAT URE | тот AL | TOTAL.MAT URE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FS | 7 |  | 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 7 |  |
| SPNG <br> FS | $\begin{aligned} & 199 \\ & 8 \end{aligned}$ | 3 | 2 | 7 | 4 | 49 | 1920 | 4685 | 1815 | 337 | 153 | 125 | 88 | 147 | 135 | 86 | 13 | 2 | 3 |  |  | 8 | 12 | 9573 | 752 |
| SPNG <br> FS | $\begin{aligned} & 199 \\ & 9 \end{aligned}$ |  | 6 | 59 | 13 | 134 | 2736 | 3010 | 193 | 106 | 83 | 109 | 143 | 390 | 645 | 402 | 69 |  |  |  |  | 8 | 14 | 8098 | 1841 |
| SPNG <br> FS | $\begin{aligned} & 200 \\ & 0 \end{aligned}$ |  | 7 | $\begin{aligned} & 372 \\ & 9 \end{aligned}$ | 2046 | 17 | 554 | 1947 | 489 | 277 | 486 | 756 | 1252 | 999 | 1021 | 199 | 34 | 13 |  |  |  | 7 | 12 | $\begin{aligned} & 1382 \\ & 7 \end{aligned}$ | 4760 |
| SPNG FS | $\begin{aligned} & 200 \\ & 1 \end{aligned}$ |  | 68 | 4 | 1 | 153 | 3241 | 5085 | 659 | 225 | 206 | 205 | 236 | 692 | 407 | 120 | 22 | 9 |  |  |  | 8 | 13 | $\begin{aligned} & 1133 \\ & 1 \end{aligned}$ | 1896 |
| SPNG <br> FS | $\begin{aligned} & 200 \\ & 2 \end{aligned}$ |  | 4 | 20 |  | 133 | 2333 | 2013 | 284 | 50 | 58 | 54 | 60 | 231 | 314 | 72 | 9 |  |  |  |  | 8 | 13 | 5634 | 798 |
| SPNG <br> FS | $\begin{aligned} & 200 \\ & 3 \end{aligned}$ |  | 4 | 950 | 567 | 4 | 77 | 221 | 57 | 39 | 28 | 16 | 22 | 17 | 23 | 16 | 5 | 1 |  |  |  | 5 | 12 | 2047 | 128 |
| SPNG <br> FS | $\begin{aligned} & 200 \\ & 4 \end{aligned}$ |  | 6 | 22 | 4 | 43 | 2289 | 3808 | 443 | 110 | 83 | 58 | 219 | 931 | 776 | 303 | 2 | 1 |  |  |  | 8 | 13 | 9097 | 2372 |
| $\begin{aligned} & \text { SPNG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 200 \\ & 5 \end{aligned}$ |  | 16 | 451 | 25 | 9 | 754 | 1007 | 207 | 85 | 102 | 30 | 54 | 257 | 218 | 90 | 44 | 2 |  |  |  | 8 | 13 | 3349 | 797 |
| SPNG <br> FS | $\begin{aligned} & 200 \\ & 6 \end{aligned}$ |  | 14 | 156 | 160 | 50 | 2238 | 8913 | 4507 | 175 | 94 | 9 | 36 | 229 | 419 | 169 | 9 | 2 |  |  |  | 7 | 14 | $\begin{aligned} & 1718 \\ & 1 \end{aligned}$ | 968 |
| SPNG FS | $\begin{aligned} & 200 \\ & 7 \end{aligned}$ |  | 49 | 40 | 1 | 111 | 3025 | 6620 | 1099 | 129 | 260 | 81 | 7 | 93 | 215 | 89 | 21 | 3 |  |  |  | 7 | 12 | $\begin{aligned} & 1184 \\ & 3 \end{aligned}$ | 768 |
| SPNG <br> FS | $\begin{aligned} & 200 \\ & 8 \end{aligned}$ | 7 | 4 | 92 | 247 | 1 | 936 | 1561 | 1326 | 234 | 1483 | 304 | 537 | 11 | 833 | 201 | 186 | 11 |  |  |  | 9 | 12 | 7974 | 3566 |
| SPNG <br> FS | $\begin{aligned} & 200 \\ & 9 \end{aligned}$ | 1 | 17 | 53 | 125 | 9 | 2582 | 3816 | 4105 | 119 | 250 | 45 | 142 | 59 | 819 | 120 | 17 | 1 | 1 |  |  | 8 | 13 | $\begin{aligned} & 1228 \\ & 3 \end{aligned}$ | 1456 |
| SPNG <br> FS | $\begin{aligned} & 201 \\ & 0 \end{aligned}$ |  | 55 | 102 | 5 | 232 | $\begin{aligned} & 1309 \\ & 0 \end{aligned}$ | $\begin{aligned} & 2203 \\ & 2 \end{aligned}$ | 3169 | 1160 | 1056 | 89 | 82 | 179 | 1007 | 1981 | 518 | 9 |  |  |  | 8 | 14 | $\begin{aligned} & 4476 \\ & 6 \end{aligned}$ | 4920 |
| SPNG <br> FS | $\begin{aligned} & 201 \\ & 1 \end{aligned}$ |  | 29 | 260 | 105 | 46 | 2805 | 5511 | 1278 | 148 | 340 | 145 | 100 | 144 | 591 | 724 | 134 | 3 | 1 |  |  | 8 | 14 | $\begin{aligned} & 1236 \\ & 4 \end{aligned}$ | 2182 |
| SPNG <br> FS | $\begin{aligned} & 201 \\ & 2 \end{aligned}$ |  | 29 | 132 | 35 | 556 | 7550 | 7844 | 1364 | 88 | 53 | 59 | 170 | 1051 | 2394 | 1553 | 432 | 21 |  |  |  | 8 | 14 | $\begin{aligned} & 2333 \\ & 1 \end{aligned}$ | 5734 |
| SPNG <br> FS | $\begin{aligned} & 201 \\ & 3 \end{aligned}$ |  |  | 2 | 11 | 126 | 2163 | 4664 | 854 | 302 | 609 | 251 | 61 | 110 | 123 | 140 | 64 | 7 |  |  |  | 8 | 12 | 9486 | 1364 |
| SPNG <br> FS | $\begin{aligned} & 201 \\ & 4 \end{aligned}$ |  | 75 | 117 | 6 | 12 | 263 | 465 | 79 | 1083 | 1175 | 1174 | 1266 | 998 | 2444 | 3623 | 817 | 31 | 1 |  |  | 12 | 13 | $\begin{aligned} & 1363 \\ & 0 \end{aligned}$ | 11530 |
| SPNG <br> FS | $\begin{aligned} & 201 \\ & 5 \end{aligned}$ |  | 13 | 67 | 3 | 58 | 1889 | 4248 | 534 | 75 | 465 | 750 | 970 | 695 | 1173 | 1473 | 453 | 70 | 1 |  |  | 10 | 13 | $\begin{aligned} & 1293 \\ & 7 \end{aligned}$ | 6050 |
| SPNG <br> FS | $\begin{aligned} & 201 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 0.1 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.8 \\ & 5 \end{aligned}$ | 0.04 | 0.39 | 9 | 24 | 4 | 9 | 7 | 3 | 6 | 5 | 6 | 2 | 0.25 | $\begin{aligned} & 0.0 \\ & 3 \end{aligned}$ |  |  |  | 9 | 12 | 77 | 29 |
| SPNG <br> FS | $\begin{aligned} & 201 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 0.1 \\ & 8 \end{aligned}$ | 0.01 | 0.14 | 6 | 18 | 7 | 1 | 2 | 3 | 4 | 6 | 10 | 9 | 2 | $\begin{aligned} & 0.1 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 3 \end{aligned}$ |  |  | 10 | 14 | 67 | 34 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 200 \\ & 1 \end{aligned}$ |  | 2 |  | 2 | 2 | 4 |  | 88 | 10 | 104 | 266 | 323 | 1334 | 2259 | 460 | 81 |  |  |  |  | 13 | 14 | 4934 | 4827 |
| SPPGF | 200 |  |  |  |  |  |  |  |  | 1 | 4 | 90 | 212 | 791 | 843 | 313 | 60 |  |  |  |  | 14 | 14 | 2314 | 2313 |


| SURVE | Yea | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | $\begin{aligned} & 2 \\ & 0 \end{aligned}$ | $\begin{gathered} \mathrm{M} \\ \mathrm{~L} \end{gathered}$ | ML.MAT URE | тот AL | TOTAL.MAT URE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| S | 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 200 \\ & 3 \end{aligned}$ |  |  |  |  |  | 1 |  | 3 | 15 | 22 | 21 | 62 | 268 | 426 | 249 | 51 | 2 | 1 |  |  | 14 | 14 | 1121 | 1102 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 200 \\ & 4 \end{aligned}$ |  | 1 |  |  |  | 5 | 2 |  | 4 | 5 | 18 | 100 | 312 | 483 | 319 | 43 | 1 |  |  |  | 14 | 14 | 1293 | 1281 |
| $\begin{aligned} & \text { SPPGF } \\ & \mathrm{S} \end{aligned}$ | $\begin{aligned} & 200 \\ & 5 \end{aligned}$ |  | 1 |  | 1 | 6 | 1 | 18 | 10 | 9 | 14 | 7 | 101 | 530 | 935 | 705 | 226 | 18 |  |  |  | 14 | 14 | 2581 | 2536 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 200 \\ & 6 \end{aligned}$ |  |  | 1 | 1 | 6 | 91 | 89 | 21 | 34 | 75 | 27 | 45 | 335 | 670 | 555 | 197 | 10 | 1 |  |  | 13 | 14 | 2158 | 1914 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 200 \\ & 7 \end{aligned}$ |  |  |  |  | 3 | 4 | 9 | 15 | 12 | 9 | 27 | 25 | 72 | 151 | 144 | 26 | 4 |  |  |  | 13 | 14 | 501 | 458 |
| $\begin{aligned} & \text { SPPGF } \\ & \mathrm{s} \end{aligned}$ | $\begin{aligned} & 200 \\ & 8 \end{aligned}$ |  | 1 |  |  |  | 1 | 13 | 7 | 16 | 13 | 55 | 106 | 237 | 457 | 302 | 78 | 5 |  |  |  | 14 | 14 | 1292 | 1254 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 200 \\ & 9 \end{aligned}$ |  | 6 | 5 |  | 2 | 7 | 8 | 1 |  | 1 | 154 | 318 | 924 | 1201 | 1172 | 324 | 7 |  |  |  | 14 | 14 | 4130 | 4101 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 201 \\ & 0 \end{aligned}$ | 1 |  |  | 1 | 5 | 14 | 3 | 1 | 5 | 2 | 31 | 284 | 521 | 717 | 459 | 123 | 10 |  |  |  | 14 | 14 | 2178 | 2148 |
| $\begin{aligned} & \text { SPPGF } \\ & \mathrm{S} \end{aligned}$ | $\begin{aligned} & 201 \\ & 1 \end{aligned}$ |  |  |  |  |  |  |  | 3 | 16 | 18 | 5 | 147 | 671 | 792 | 429 | 122 | 13 |  | 2 |  | 14 | 14 | 2220 | 2200 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 201 \\ & 2 \end{aligned}$ |  |  |  | 1 | 1 |  |  | 2 | 2 | 1 | 8 | 70 | 369 | 468 | 218 | 66 | 3 |  |  |  | 14 | 14 | 1208 | 1202 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 201 \\ & 3 \end{aligned}$ |  |  |  | 1 |  | 7 | 22 | 6 | 9 |  | 1 | 42 | 435 | 889 | 480 | 141 | 12 | 1 |  |  | 14 | 14 | 2045 | 2000 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 201 \\ & 4 \end{aligned}$ |  | 10 | 9 |  | 1 |  | 3 | 17 | 62 | 11 | 6 | 85 | 2453 | 6703 | 3168 | 2115 | 162 | 82 |  |  | 14 | 14 | $\begin{aligned} & 1488 \\ & 9 \end{aligned}$ | 14787 |
| $\begin{aligned} & \text { SPPGF } \\ & \mathrm{s} \end{aligned}$ | $\begin{aligned} & 201 \\ & 5 \end{aligned}$ |  |  |  | 2 | 1 |  |  | 1 | 1 |  |  | 32 | 300 | 471 | 316 | 151 | 43 |  |  |  | 14 | 14 | 1318 | 1313 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 201 \\ & 6 \end{aligned}$ |  |  | $\begin{aligned} & 0.0 \\ & 4 \end{aligned}$ |  |  |  | 0.02 |  | 0.16 | 0.06 |  | 0.1 | 2 | 4 | 3 | 1 | $\begin{aligned} & 0.2 \\ & 5 \end{aligned}$ |  |  |  | 14 | 14 | 11 | 11 |
| $\begin{aligned} & \text { SPPGF } \\ & \text { S } \end{aligned}$ | $\begin{aligned} & 201 \\ & 7 \end{aligned}$ |  | 1 | $\begin{aligned} & 0.3 \\ & 5 \end{aligned}$ |  |  |  | 0.2 |  |  | 0.02 | 0.35 | 0.52 | 3 | 10 | 10 | 5 | $\begin{aligned} & 0.3 \\ & 3 \end{aligned}$ |  |  |  | 14 | 15 | 31 | 29 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 198 \\ & 6 \end{aligned}$ |  |  |  |  |  |  |  | 0.5 |  |  |  |  |  |  |  |  |  |  |  |  | 8 |  |  |  |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 198 \\ & 7 \end{aligned}$ |  |  |  |  |  |  |  | 0.5 | 0.5 | 2 | 0.5 |  |  |  |  |  |  |  |  |  | 10 | 10 | 4 | 2 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 198 \\ & 8 \end{aligned}$ |  |  |  | 0.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 4 |  |  |  |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 198 \\ & 9 \end{aligned}$ |  |  |  |  |  |  | 0.5 |  |  |  |  |  |  |  |  |  |  |  |  |  | 7 |  |  |  |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 0 \end{aligned}$ |  |  |  | 1 |  | 0.5 | 1 | 2 | 24 | 54 | 50 | 43 | 12 | 1 |  |  |  |  |  |  | 11 | 11 | 188 | 160 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 1 \end{aligned}$ |  |  |  |  |  | 1 | 0.5 | 8 | 38 | 183 | 266 | 316 | 48 | 16 |  |  |  |  |  |  | 11 | 11 | 876 | 829 |
| WCSG | 199 |  |  |  |  |  | 1 |  | 10 | 38 | 468 | 1145 | 4001 | 1626 | 486 |  |  |  |  |  |  | 12 | 12 | 7775 | 7726 |


| SURVE $\mathrm{Y}$ | $\begin{gathered} \text { YEA } \\ \mathrm{R} \end{gathered}$ | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | $\begin{aligned} & 2 \\ & 0 \end{aligned}$ | $\begin{gathered} \mathrm{M} \\ \mathrm{~L} \end{gathered}$ | ML.MAT URE | тот <br> AL | TOTAL.MAT URE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FS | 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 3 \end{aligned}$ |  |  |  |  |  |  | 4 |  | 2 | 9 | 60 | 155 | 72 | 16 |  | 0.5 |  |  |  |  | 12 | 12 | 319 | 312 |
| WCSG <br> FS | $\begin{aligned} & 199 \\ & 4 \end{aligned}$ |  |  |  |  |  |  |  |  | 0.5 | 0.5 | 0.5 |  |  | 0.5 |  |  |  |  |  |  | 11 | 12 | 2 | 2 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 5 \end{aligned}$ |  |  |  |  |  |  |  |  | 8 | 36 | 194 | 294 | 398 | 199 | 22 |  |  |  |  |  | 12 | 12 | 1150 | 1142 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 6 \end{aligned}$ |  |  |  | 2 |  | 4 | 3 |  |  |  | 1 | 55 | 610 | 1574 | 304 |  |  |  |  |  | 14 | 14 | 2552 | 2544 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 7 \end{aligned}$ |  |  | 4 |  |  | 0.5 | 6 | 9 | 4 | 6 | 25 | 108 | 203 | 157 | 40 | 4 |  |  |  |  | 13 | 13 | 568 | 544 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 8 \end{aligned}$ |  |  |  | 1 |  | 1 | 5 | 2 |  | 1 | 2 |  | 3 |  |  |  |  |  |  |  | 9 | 12 | 15 | 6 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 199 \\ & 9 \end{aligned}$ |  |  | 1 |  |  | 2 | 5 | 1 | 1 |  | 1 | 2 | 1 |  |  |  |  |  |  |  | 8 | 12 | 14 | 4 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 200 \\ & 0 \end{aligned}$ |  |  |  |  |  |  | 2 | 2 | 39 | 110 | 216 | 288 | 182 | 92 | 46 | 6 |  |  |  |  | 12 | 12 | 983 | 940 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 200 \\ & 1 \end{aligned}$ |  | 1 |  |  |  |  |  | 1 | 4 | 15 | 28 | 59 | 134 | 240 | 103 | 10 | 4 |  |  |  | 14 | 14 | 599 | 593 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 200 \\ & 2 \end{aligned}$ |  |  |  |  |  | 1 | 8 | 2 | 1 | 82 | 742 | 3211 | 5601 | 5772 | 1497 | 167 | 1 |  |  |  | 13 | 13 | $\begin{aligned} & 1708 \\ & 4 \end{aligned}$ | 17072 |
| WCSG FS | $\begin{aligned} & 200 \\ & 3 \end{aligned}$ |  |  | 1 |  |  |  | 3 | 52 |  | 53 | 281 | 1473 | 3066 | 4895 | 3083 | 309 | 28 |  |  |  | 14 | 14 | $\begin{aligned} & 1324 \\ & 4 \end{aligned}$ | 13188 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 200 \\ & 4 \end{aligned}$ |  |  |  | 1 |  |  | 2 | 2 | 43 | 82 | 743 | 4569 | 8600 | 9514 | 5692 | 948 | 84 |  |  |  | 14 | 14 | $\begin{aligned} & 3028 \\ & 0 \end{aligned}$ | 30232 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 200 \\ & 5 \end{aligned}$ |  | 2 |  |  |  |  | 24 | 3 | 23 | 25 | 110 | 435 | 1085 | 1708 | 792 | 130 | 6 |  |  |  | 14 | 14 | 4343 | 4291 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 200 \\ & 6 \end{aligned}$ |  | 1 | 2 | 1 |  | 1 | 4 |  | 10 | 218 | 232 | 452 | 1396 | 2852 | 2051 | 434 | 72 |  |  |  | 14 | 14 | 7726 | 7706 |
| WCSG <br> FS | $\begin{aligned} & 200 \\ & 7 \end{aligned}$ |  |  | 2 | 2 |  | 2 | 1 | 3 | 21 | 159 | 780 | 2923 | 5194 | 6888 | 5283 | 1523 | 116 |  |  |  | 14 | 14 | $\begin{aligned} & 2289 \\ & 7 \end{aligned}$ | 22866 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 200 \\ & 8 \end{aligned}$ |  | 1 | 1 |  |  | 16 | 37 | 36 | 187 | 468 | 1395 | 3213 | 9893 | $\begin{aligned} & 2275 \\ & 8 \end{aligned}$ | $\begin{aligned} & 1839 \\ & 9 \end{aligned}$ | 6288 | 575 | 71 |  |  | 14 | 14 | $\begin{aligned} & 6333 \\ & 8 \end{aligned}$ | 63060 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 200 \\ & 9 \end{aligned}$ |  |  | 1 |  |  | 1 |  | 4 | 52 | 2442 | 2093 | 440 | 331 | 287 | 246 | 129 | 10 |  |  |  | 11 | 11 | 6038 | 5978 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 201 \\ & 0 \end{aligned}$ |  |  |  |  |  |  |  |  |  |  | 530 | 1443 | 1384 | 1357 | 828 | 149 | 29 |  |  |  | 13 | 13 | 5720 | 5720 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 201 \\ & 1 \end{aligned}$ |  | 1 | 4 | 1 |  | 1 | 5 | 254 | 1015 | 2034 | 7613 | $\begin{aligned} & 1891 \\ & 8 \end{aligned}$ | $\begin{aligned} & 1447 \\ & 8 \end{aligned}$ | 6445 | 2006 | 236 | 23 |  |  |  | 12 | 12 | $\begin{aligned} & 5303 \\ & 4 \end{aligned}$ | 51753 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 201 \\ & 2 \end{aligned}$ |  |  | 1 |  |  | 1 | 2 |  | 103 | 9 | 1267 | 6545 | $\begin{aligned} & 2633 \\ & 7 \end{aligned}$ | $\begin{aligned} & 2936 \\ & 1 \end{aligned}$ | $\begin{aligned} & 2733 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1585 \\ & 7 \end{aligned}$ | $\begin{aligned} & 150 \\ & 5 \end{aligned}$ | $\begin{gathered} 49 \\ 6 \end{gathered}$ |  |  | 14 | 14 | $\begin{aligned} & 1088 \\ & 17 \end{aligned}$ | 108710 |
| $\begin{aligned} & \text { WCSG } \\ & \text { FS } \end{aligned}$ | $\begin{aligned} & 201 \\ & 3 \end{aligned}$ |  |  |  | 1 |  |  | 1 |  |  | 1 | 143 | 3201 | $\begin{aligned} & 1528 \\ & 2 \end{aligned}$ | $\begin{aligned} & 1128 \\ & 8 \end{aligned}$ | 3934 | 858 | 6 | 1 |  |  | 14 | 14 | $\begin{aligned} & 3471 \\ & 6 \end{aligned}$ | 34714 |
| WCSG | 201 |  | 48 | 457 | 386 | 48 | 3 | 7 | 63 | 21 | 98 | 876 | 1166 | 3026 | 3923 | 1093 | 1363 | 111 | 1 |  |  | 13 | 14 | 9558 | 94553 |



Table 3.6.1.1. Boarfish in ICES Subareas 27.6, 7, 8. IBTS length-frequency data converted to age-structured index by application of the 2010 common ALK rounded down to 1 cm length classes

| Sur vey | $\begin{aligned} & \mathrm{Ye} \\ & \text { ar } \end{aligned}$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| EVH | 19 | 23 | 187 | 600 | 374 | 391 | 393 | 706 | 586 | 421 | 483 | 425 | 146 | 242 | 169 | 121 | 62 | 121 | 15 | 659 | 62 | 848 | 768 | 21 | 32 | 54 | 10 | 15 | 51 | 31 | 41 |
| OE | 97 |  | 6 | 3 | 1 | 1 | 8 | 5 | 7 | 8 | 2 | 9 | 1 | 8 | 9 | 4 | 3 | 5 | 9 |  | 3 |  |  | 4 | 5 | 3 | 0 | 8 |  | 4 | 6 |
| EVH | 19 | 31 | 129 | 159 | 624 | 624 | 559 | 743 | 573 | 377 | 480 | 438 | 146 | 284 | 163 | 161 | 67 | 122 | 23 | 904 | 67 | 965 | 104 | 32 | 47 | 75 | 18 | 23 | 93 | 46 | 35 |
| OE | 98 |  | 77 | 97 | 8 | 7 | 1 | 5 | 2 | 7 | 6 | 6 | 3 | 3 | 5 | 9 | 6 | 4 | 2 |  | 6 |  | 2 | 7 | 6 | 2 | 7 | 1 |  | 1 | 3 |
| EVH | 19 | 65 | 757 | 312 | 199 | 873 | 349 | 330 | 271 | 190 | 272 | 235 | 743 | 154 | 975 | 893 | 28 | 647 | 62 | 474 | 28 | 477 | 509 | 91 | 24 | 31 | 53 | 61 | 27 | 12 | 19 |
| OE | 99 |  | 6 | 23 | 15 | 2 | 9 | 8 | 5 | 5 | 0 | 7 |  | 0 |  |  | 5 |  |  |  | 5 |  |  |  | 6 | 7 |  |  |  | 3 | 7 |
| EVH | 20 | 21 | 176 | 277 | 125 | 179 | 155 | 187 | 142 | 973 | 110 | 949 | 320 | 516 | 379 | 255 | 12 | 260 | 25 | 138 | 12 | 178 | 153 | 37 | 71 | 10 | 19 | 24 | 99 | 49 | 92 |
| OE | 00 | 7 | 76 | 30 | 86 | 86 | 25 | 40 | 97 | 7 | 41 | 0 | 8 | 0 | 7 | 6 | 66 | 4 | 3 | 4 | 66 | 2 | 8 | 4 | 4 | 22 | 8 | 5 |  | 1 | 1 |
| EVH | 20 | 73 | 143 | 413 | 203 | 254 | 219 | 162 | 924 | 452 | 454 | 395 | 133 | 205 | 132 | 109 | 57 | 959 | 15 | 684 | 57 | 780 | 710 | 30 | 45 | 50 | 25 | 14 | 12 | 29 | 30 |
| OE | 01 | 3 | 89 | 13 | 57 | 67 | 21 | 11 | 7 | 5 | 3 | 1 | 2 | 7 | 2 | 8 | 8 |  | 3 |  | 8 |  |  | 4 | 6 | 8 | 4 | 7 | 9 | 0 | 6 |
| EVH | 20 | 43 | 671 | 317 | 184 | 127 | 838 | 711 | 476 | 285 | 342 | 301 | 994 | 180 | 112 | 100 | 42 | 796 | 11 | 573 | 42 | 617 | 625 | 19 | 32 | 42 | 12 | 11 | 65 | 22 | 24 |
| OE | 02 |  | 9 | 28 | 55 | 84 | 9 | 5 | 7 | 1 | 9 | 8 |  | 6 | 3 | 9 | 1 |  | 7 |  | 1 |  |  | 2 | 4 | 9 | 8 | 3 |  | 7 | 4 |
| EVH | 20 | 64 | 509 | 399 | 734 | 183 | 172 | 161 | 107 | 627 | 762 | 685 | 226 | 429 | 250 | 245 | 10 | 183 | 32 | 138 | 10 | 146 | 155 | 49 | 76 | 11 | 31 | 32 | 15 | 64 | 53 |
| OE | 03 |  |  | 3 | 8 | 71 | 76 | 13 | 98 | 0 | 0 | 2 | 7 | 4 | 1 | 6 | 09 | 8 | 6 | 7 | 09 | 2 | 7 | 1 | 3 | 04 | 0 | 2 | 5 | 4 | 2 |
| EVH | 20 | 54 | 126 | 197 | 126 | 172 | 222 | 412 | 322 | 206 | 287 | 305 | 106 | 242 | 939 | 150 | 90 | 917 | 38 | 114 | 90 | 110 | 116 | 81 | 92 | 96 | 72 | 36 | 36 | 71 | 18 |
| OE | 04 | 5 | 5 | 6 | 1 | 2 | 7 | 4 | 8 | 1 | 1 | 8 | 6 | 6 |  | 9 | 1 |  | 2 | 2 | 1 | 0 | 0 | 7 | 5 | 2 | 6 | 0 | 6 | 5 | 1 |
| EVH | 20 | 10 | 210 | 260 | 149 | 209 | 301 | 716 | 599 | 417 | 530 | 487 | 164 | 314 | 179 | 177 | 83 | 136 | 28 | 106 | 83 | 114 | 118 | 48 | 63 | 87 | 33 | 30 | 20 | 54 | 39 |
| OE | 05 | 70 | 2 | 3 | 7 | 8 | 5 | 0 | 2 | 7 | 1 | 3 | 2 | 4 | 6 | 6 | 3 | 8 | 5 | 5 | 3 | 0 | 4 | 6 | 9 | 7 | 2 | 8 | 1 | 6 | 4 |
| EVH | 20 | 21 | 358 | 265 | 480 | 219 | 138 | 148 | 133 | 947 | 152 | 148 | 485 | 117 | 557 | 725 | 31 | 445 | 12 | 464 | 31 | 434 | 496 | 24 | 30 | 37 | 18 | 11 | 93 | 24 | 10 |
| OE | 06 | 7 | 34 | 93 | 3 | 9 | 6 | 9 | 2 |  | 1 | 4 |  | 0 |  |  | 1 |  | 5 |  | 1 |  |  | 5 | 8 | 3 | 4 | 6 |  | 2 | 3 |
| EVH | 20 | 66 | 168 | 122 | 653 | 169 | 491 | 431 | 296 | 171 | 245 | 239 | 788 | 180 | 820 | 112 | 48 | 678 | 20 | 715 | 48 | 668 | 778 | 38 | 46 | 59 | 28 | 19 | 14 | 38 | 15 |
| OE | 07 | 1 | 18 | 140 | 69 | 86 | 9 | 6 | 7 | 5 | 2 | 2 |  | 2 |  | 4 | 4 |  | 4 |  | 4 |  |  | 1 | 7 | 4 | 2 | 8 | 6 | 5 | 0 |
| EVH | 20 | 32 | 416 | 258 | 168 | 134 | 771 | 377 | 187 | 827 | 913 | 818 | 266 | 486 | 245 | 299 | 12 | 187 | 49 | 191 | 12 | 176 | 206 | 10 | 12 | 15 | 69 | 42 | 35 | 83 | 46 |
| OE | 08 | 44 | 11 | 758 | 378 | 061 | 06 | 38 | 50 | 7 | 2 | 3 | 0 | 8 | 8 | 2 | 26 | 6 | 2 | 9 | 26 | 5 | 2 | 64 | 37 | 23 | 8 | 0 | 2 | 5 | 0 |
| EVH | 20 | 32 | 133 | 368 | 121 | 562 | 598 | 778 | 544 | 305 | 444 | 423 | 136 | 307 | 138 | 196 | 61 | 111 | 30 | 106 | 61 | 956 | 129 | 39 | 49 | 95 | 15 | 30 | 78 | 61 | 23 |
| OE | 09 | 7 | 38 | 29 | 94 | 6 | 2 | 8 | 3 | 4 | 3 | 0 | 4 | 9 | 2 | 5 | 8 | 4 | 9 | 4 | 8 |  | 5 | 8 | 3 | 7 | 5 | 6 |  | 1 | 5 |
| EVH | 20 | 66 | 336 | 839 | 350 | 216 | 235 | 342 | 230 | 126 | 163 | 145 | 464 | 900 | 471 | 555 | 16 | 345 | 69 | 295 | 16 | 274 | 349 | 92 | 13 | 24 | 31 | 66 | 16 | 13 | 86 |
| OE | 10 | 6 | 01 | 03 | 48 | 78 | 03 | 10 | 37 | 43 | 03 | 19 | 7 | 8 | 6 | 1 | 89 | 7 | 0 | 7 | 89 | 5 | 0 | 0 | 68 | 35 | 2 | 9 | 0 | 31 | 8 |
| EVH | 20 | 37 | 221 | 124 | 149 | 287 | 261 | 318 | 239 | 155 | 194 | 169 | 554 | 101 | 653 | 566 | 22 | 451 | 59 | 319 | 22 | 340 | 348 | 10 | 17 | 23 | 61 | 61 | 38 | 11 | 14 |
| OE | 11 | 0 | 2 | 71 | 82 | 29 | 14 | 44 | 15 | 35 | 73 | 64 | 2 | 76 | 4 | 3 | 62 | 3 | 7 | 7 | 62 | 8 | 6 | 77 | 62 | 39 | 6 | 9 | 8 | 26 | 14 |
| EVH | 20 | 73 | 200 | 343 | 115 | 110 | 107 | 149 | 133 | 900 | 156 | 147 | 459 | 114 | 554 | 732 | 23 | 414 | 92 | 416 | 23 | 370 | 459 | 14 | 23 | 32 | 97 | 90 | 49 | 18 | 92 |
| OE | 12 | 8 | 89 | 48 | 35 | 98 | 95 | 79 | 08 | 4 | 62 | 14 | 8 | 67 | 0 | 5 | 25 | 2 | 0 | 4 | 25 | 3 | 5 | 47 | 56 | 18 | 9 | 8 | 0 | 15 | 8 |
| EVH | 20 | 14 | 164 | 369 | 380 | 103 | 920 | 113 | 112 | 829 | 144 | 137 | 437 | 109 | 536 | 689 | 25 | 406 | 98 | 420 | 25 | 381 | 449 | 18 | 26 | 32 | 13 | 91 | 69 | 18 | 94 |
| OE | 13 | 2 | 7 | 5 | 5 | 88 | 7 | 85 | 71 | 9 | 85 | 97 | 4 | 61 | 4 | 3 | 50 | 8 | 1 | 5 | 50 | 6 | 4 | 72 | 50 | 28 | 84 | 4 | 2 | 30 | 4 |
| EVH | 20 | 20 | 152 | 236 | 380 | 129 | 173 | 276 | 249 | 174 | 274 | 250 | 791 | 182 | 991 | 111 | 34 | 710 | 12 | 597 | 34 | 564 | 681 | 16 | 29 | 46 | 78 | 14 | 60 | 24 | 18 |
| OE | 14 | 81 | 4 | 5 | 5 | 88 | 15 | 92 | 54 | 60 | 10 | 16 | 1 | 66 | 8 | 60 | 65 | 7 | 27 | 7 | 65 | 4 | 3 | 36 | 61 | 34 | 2 | 38 | 7 | 43 | 53 |
| EVH | 20 | 60 | 192 | 175 | 108 | 358 | 176 | 331 | 267 | 174 | 255 | 228 | 720 | 153 | 839 | 944 | 30 | 595 | 10 | 532 | 30 | 495 | 580 | 17 | 29 | 39 | 10 | 11 | 76 | 19 | 15 |
| OE | 15 | 85 | 33 | 572 | 367 | 91 | 18 | 96 | 70 | 33 | 62 | 40 | 8 | 96 | 6 | 5 | 78 | 2 | 33 | 5 | 78 | 0 | 9 | 44 | 69 | 37 | 97 | 93 | 3 | 65 | 51 |
| EVH | 20 | 12 | 736 | 210 | 183 | 329 | 286 | 436 | 415 | 302 | 497 | 454 | 142 | 336 | 179 | 208 | 66 | 128 | 23 | 117 | 66 | 107 | 128 | 39 | 64 | 87 | 23 | 22 | 11 | 44 | 32 |
| OE | 16 | 56 | 0 | 28 | 55 | 37 | 79 | 26 | 81 | 74 | 97 | 44 | 38 | 54 | 99 | 15 | 33 | 39 | 42 | 04 | 33 | 34 | 85 | 10 | 23 | 85 | 22 | 19 | 74 | 13 | 66 |


| $\begin{aligned} & \text { Sur } \\ & \text { vey } \end{aligned}$ | $\begin{aligned} & \mathrm{Ye} \\ & \mathrm{ar} \end{aligned}$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| EVH OE | $\begin{aligned} & 20 \\ & 17 \end{aligned}$ | $23$ | 187 | 263 | 50 | 0.91 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| IGFS | $\begin{aligned} & 20 \\ & 03 \end{aligned}$ | 55 | 126 | 517 | 930 | $\begin{aligned} & 230 \\ & 6 \end{aligned}$ | $\begin{aligned} & 185 \\ & 8 \end{aligned}$ | $\begin{aligned} & 143 \\ & 3 \end{aligned}$ | $\begin{aligned} & 124 \\ & 4 \end{aligned}$ | 842 | $154$ | $\begin{aligned} & 154 \\ & 5 \end{aligned}$ | 494 | $\begin{aligned} & 130 \\ & 9 \end{aligned}$ | 576 | 842 | $\begin{aligned} & 31 \\ & 7 \end{aligned}$ | 467 | $\begin{aligned} & 14 \\ & 8 \end{aligned}$ | 527 | $\begin{aligned} & 31 \\ & 7 \end{aligned}$ | 461 | 585 | $\begin{aligned} & 28 \\ & 7 \end{aligned}$ | $\begin{aligned} & 32 \\ & 4 \end{aligned}$ | $\begin{aligned} & 44 \\ & 1 \end{aligned}$ | $\begin{aligned} & 17 \\ & 9 \end{aligned}$ | $\begin{aligned} & 15 \\ & 1 \end{aligned}$ | $\begin{aligned} & 10 \\ & 9 \end{aligned}$ | $\begin{aligned} & 26 \\ & 3 \end{aligned}$ | 96 |
| IGFS | $\begin{aligned} & 20 \\ & 04 \end{aligned}$ | $\begin{aligned} & 12 \\ & 0 \end{aligned}$ | 418 | $\begin{aligned} & 142 \\ & 2 \end{aligned}$ | 594 | 396 | 484 | $\begin{aligned} & 130 \\ & 3 \end{aligned}$ | $\begin{aligned} & 134 \\ & 1 \end{aligned}$ | 993 | ${ }_{3}^{171}$ | ${ }_{3}^{177}$ | 589 | $\begin{aligned} & 149 \\ & 1 \end{aligned}$ | 618 | 948 | $\begin{aligned} & 39 \\ & 0 \end{aligned}$ | 543 | $\begin{aligned} & 18 \\ & 9 \end{aligned}$ | 584 | $\begin{aligned} & 39 \\ & 0 \end{aligned}$ | 537 | 672 | $\begin{aligned} & 31 \\ & 7 \end{aligned}$ | $\begin{aligned} & 35 \\ & 0 \end{aligned}$ | $\begin{aligned} & 52 \\ & 5 \end{aligned}$ | $\begin{aligned} & 20 \\ & 3 \end{aligned}$ | $\begin{aligned} & 18 \\ & 1 \end{aligned}$ | $\begin{aligned} & 10 \\ & 3 \end{aligned}$ | $\begin{aligned} & 36 \\ & 2 \end{aligned}$ | $\begin{aligned} & 10 \\ & 8 \end{aligned}$ |
| IGFS | $\begin{aligned} & 20 \\ & 05 \end{aligned}$ | $\begin{aligned} & 11 \\ & 9 \end{aligned}$ | 814 | 982 | 379 | 542 | 665 | ${ }_{2}^{230}$ | $\begin{aligned} & 288 \\ & 4 \end{aligned}$ | $\begin{aligned} & 236 \\ & 4 \end{aligned}$ | $\begin{aligned} & 412 \\ & 9 \end{aligned}$ | $\begin{aligned} & 414 \\ & 0 \end{aligned}$ | $\begin{aligned} & 136 \\ & 0 \end{aligned}$ | $\begin{aligned} & 343 \\ & 1 \end{aligned}$ | $\begin{aligned} & 156 \\ & 9 \end{aligned}$ | $\begin{aligned} & 214 \\ & 2 \end{aligned}$ | $\begin{aligned} & 82 \\ & 2 \end{aligned}$ | $\begin{aligned} & 128 \\ & 9 \end{aligned}$ | $\begin{aligned} & 40 \\ & 0 \end{aligned}$ | $\begin{aligned} & 128 \\ & 3 \end{aligned}$ | $\begin{aligned} & 82 \\ & 2 \end{aligned}$ | $\begin{aligned} & 117 \\ & 7 \end{aligned}$ | $\begin{aligned} & 150 \\ & 9 \end{aligned}$ | $\begin{aligned} & 68 \\ & 9 \end{aligned}$ | $\begin{aligned} & 70 \\ & 3 \end{aligned}$ | $\begin{aligned} & 11 \\ & 54 \end{aligned}$ | $\begin{aligned} & 34 \\ & 9 \end{aligned}$ | $\begin{aligned} & 36 \\ & 3 \end{aligned}$ | $\begin{aligned} & 17 \\ & 5 \end{aligned}$ | $\begin{aligned} & 72 \\ & 4 \end{aligned}$ | 28 6 |
| IGFS | $\begin{aligned} & 20 \\ & 06 \end{aligned}$ | $\begin{aligned} & 17 \\ & 6 \end{aligned}$ | 849 | $\begin{aligned} & 157 \\ & 2 \end{aligned}$ | $\begin{aligned} & 198 \\ & 8 \end{aligned}$ | $\begin{aligned} & 471 \\ & 9 \end{aligned}$ | $\begin{aligned} & 505 \\ & 1 \end{aligned}$ | $\begin{aligned} & 688 \\ & 5 \end{aligned}$ | $\begin{aligned} & 752 \\ & 2 \end{aligned}$ | $\begin{aligned} & 517 \\ & 9 \end{aligned}$ | $\begin{aligned} & 121 \\ & 77 \end{aligned}$ | $\begin{aligned} & 130 \\ & 18 \end{aligned}$ | $\begin{aligned} & 415 \\ & 1 \end{aligned}$ | $\begin{aligned} & 121 \\ & 78 \end{aligned}$ | $\begin{aligned} & 444 \\ & 8 \end{aligned}$ | $\begin{aligned} & 818 \\ & 9 \end{aligned}$ | $\begin{aligned} & 32 \\ & 97 \end{aligned}$ | $\begin{aligned} & 398 \\ & 9 \end{aligned}$ | $\begin{aligned} & 17 \\ & 08 \end{aligned}$ | $\begin{aligned} & 557 \\ & 0 \end{aligned}$ | $\begin{aligned} & 32 \\ & 97 \end{aligned}$ | $\begin{aligned} & 461 \\ & 3 \end{aligned}$ | $\begin{aligned} & 604 \\ & 8 \end{aligned}$ | $\begin{aligned} & 36 \\ & 73 \end{aligned}$ | $\begin{aligned} & 37 \\ & 75 \end{aligned}$ | $\begin{aligned} & 47 \\ & 31 \end{aligned}$ | $\begin{aligned} & 24 \\ & 59 \end{aligned}$ | $\begin{aligned} & 17 \\ & 28 \end{aligned}$ | $\begin{aligned} & 14 \\ & 96 \end{aligned}$ | $\begin{aligned} & 29 \\ & 24 \end{aligned}$ | 60 5 |
| IGFS | $\begin{aligned} & 20 \\ & 07 \end{aligned}$ | 68 | $\begin{aligned} & 105 \\ & 2 \end{aligned}$ | $\begin{aligned} & 186 \\ & 6 \end{aligned}$ | $\begin{aligned} & 138 \\ & 5 \end{aligned}$ | $\begin{aligned} & 160 \\ & 5 \end{aligned}$ | $\begin{aligned} & 164 \\ & 8 \end{aligned}$ | $\begin{aligned} & 262 \\ & 5 \end{aligned}$ | $\begin{aligned} & 262 \\ & 8 \end{aligned}$ | $\begin{aligned} & 185 \\ & 5 \end{aligned}$ | $\begin{aligned} & 354 \\ & 7 \end{aligned}$ | $\begin{aligned} & 357 \\ & 7 \end{aligned}$ | $\begin{aligned} & 114 \\ & 5 \end{aligned}$ | $\begin{aligned} & 305 \\ & 9 \end{aligned}$ | $\begin{aligned} & 129 \\ & 2 \end{aligned}$ | $\begin{aligned} & 198 \\ & 7 \end{aligned}$ | $\begin{aligned} & 72 \\ & 3 \end{aligned}$ | ${ }_{2}^{107}$ | $\begin{aligned} & 33 \\ & 2 \end{aligned}$ | $\begin{aligned} & 119 \\ & 6 \end{aligned}$ | $\begin{aligned} & 72 \\ & 3 \end{aligned}$ | $\begin{aligned} & 105 \\ & 8 \end{aligned}$ | $\begin{aligned} & 133 \\ & 5 \end{aligned}$ | $\begin{aligned} & 55 \\ & 3 \end{aligned}$ | $\begin{aligned} & 72 \\ & 2 \end{aligned}$ | $\begin{aligned} & 99 \\ & 9 \end{aligned}$ | 38 7 | $\begin{aligned} & 32 \\ & 2 \end{aligned}$ | $\begin{aligned} & 19 \\ & 3 \end{aligned}$ | $\begin{aligned} & 64 \\ & 5 \end{aligned}$ | 20 7 |
| IGFS | $\begin{aligned} & 20 \\ & 08 \end{aligned}$ | 44 | 588 | $\begin{aligned} & 170 \\ & 9 \end{aligned}$ | $\begin{aligned} & 344 \\ & 5 \end{aligned}$ | $\begin{aligned} & 123 \\ & 63 \end{aligned}$ | $\begin{aligned} & 125 \\ & 97 \end{aligned}$ | $\begin{aligned} & 132 \\ & 66 \end{aligned}$ | $\begin{aligned} & 921 \\ & 9 \end{aligned}$ | $\begin{aligned} & 522 \\ & 7 \end{aligned}$ | $\begin{aligned} & 777 \\ & 3 \end{aligned}$ | $\begin{aligned} & 779 \\ & 7 \end{aligned}$ | $\begin{aligned} & 257 \\ & 6 \end{aligned}$ | $\begin{aligned} & 606 \\ & 9 \end{aligned}$ | $\begin{aligned} & 249 \\ & 1 \end{aligned}$ | $\begin{aligned} & 388 \\ & 6 \end{aligned}$ | $\begin{aligned} & 20 \\ & 28 \end{aligned}$ | $\begin{aligned} & 218 \\ & 3 \end{aligned}$ | $\begin{aligned} & 90 \\ & 0 \end{aligned}$ | $\begin{aligned} & 299 \\ & 6 \end{aligned}$ | $\begin{aligned} & 20 \\ & 28 \end{aligned}$ | $\begin{aligned} & 263 \\ & 7 \end{aligned}$ | $\begin{aligned} & 301 \\ & 7 \end{aligned}$ | $\begin{aligned} & 23 \\ & 03 \end{aligned}$ | $\begin{aligned} & 23 \\ & 67 \end{aligned}$ | $\begin{aligned} & 24 \\ & 08 \end{aligned}$ | $\begin{aligned} & 17 \\ & 58 \end{aligned}$ | $\begin{aligned} & 76 \\ & 3 \end{aligned}$ | $\begin{aligned} & 91 \\ & 7 \end{aligned}$ | $\begin{aligned} & 14 \\ & 51 \end{aligned}$ | 42 4 |
| IGFS | $\begin{aligned} & 20 \\ & 09 \end{aligned}$ | $\begin{aligned} & 15 \\ & 8 \end{aligned}$ | 267 | 776 | $\begin{aligned} & 107 \\ & 7 \end{aligned}$ | $\begin{aligned} & 317 \\ & 4 \end{aligned}$ | $\begin{aligned} & 454 \\ & 3 \end{aligned}$ | $\begin{aligned} & 551 \\ & 3 \end{aligned}$ | $\begin{aligned} & 362 \\ & 0 \end{aligned}$ | $\begin{aligned} & 183 \\ & 9 \end{aligned}$ | $\begin{aligned} & 270 \\ & 1 \end{aligned}$ | $\begin{aligned} & 270 \\ & 6 \end{aligned}$ | 886 | $\begin{aligned} & 210 \\ & 1 \end{aligned}$ | 818 | $\begin{aligned} & 137 \\ & 3 \end{aligned}$ | $\begin{aligned} & 49 \\ & 1 \end{aligned}$ | 727 | $\begin{aligned} & 26 \\ & 1 \end{aligned}$ | 802 | $\begin{aligned} & 49 \\ & 1 \end{aligned}$ | 707 | 954 | 39 0 | $\begin{aligned} & 43 \\ & 3 \end{aligned}$ | $\begin{aligned} & 73 \\ & 8 \end{aligned}$ | $\begin{aligned} & 21 \\ & 7 \end{aligned}$ | $\begin{aligned} & 25 \\ & 5 \end{aligned}$ | $\begin{aligned} & 10 \\ & 9 \end{aligned}$ | $\begin{aligned} & 50 \\ & 8 \end{aligned}$ | 12 8 |
| IGFS | $\begin{aligned} & 20 \\ & 10 \end{aligned}$ | 51 | 374 | 747 | 902 | $\begin{aligned} & 302 \\ & 1 \end{aligned}$ | $\begin{aligned} & 659 \\ & 0 \end{aligned}$ | $\begin{aligned} & 172 \\ & 50 \end{aligned}$ | $\begin{aligned} & 132 \\ & 58 \end{aligned}$ | $\begin{aligned} & 863 \\ & 0 \end{aligned}$ | $\begin{aligned} & 100 \\ & 98 \end{aligned}$ | $\begin{aligned} & 892 \\ & 4 \end{aligned}$ | $\begin{aligned} & 300 \\ & 2 \end{aligned}$ | $\begin{aligned} & 505 \\ & 3 \end{aligned}$ | $\begin{aligned} & 315 \\ & 0 \end{aligned}$ | $\begin{aligned} & 275 \\ & 0 \end{aligned}$ | $\begin{aligned} & 12 \\ & 84 \end{aligned}$ | $\begin{aligned} & 230 \\ & 3 \end{aligned}$ | $\begin{aligned} & 41 \\ & 4 \end{aligned}$ | $\begin{aligned} & 161 \\ & 6 \end{aligned}$ | $\begin{aligned} & 12 \\ & 84 \end{aligned}$ | ${ }_{6}^{178}$ | $\begin{aligned} & 183 \\ & 2 \end{aligned}$ | $\begin{aligned} & 74 \\ & 2 \end{aligned}$ | $\begin{aligned} & 89 \\ & 7 \end{aligned}$ | $\begin{aligned} & 13 \\ & 31 \end{aligned}$ | $\begin{aligned} & 39 \\ & 5 \end{aligned}$ | $\begin{aligned} & 37 \\ & 1 \end{aligned}$ | $\begin{aligned} & 19 \\ & 7 \end{aligned}$ | $\begin{aligned} & 74 \\ & 2 \end{aligned}$ | 71 5 |
| IGFS | $\begin{aligned} & 20 \\ & 11 \end{aligned}$ | 25 | 641 | 951 | 598 | $\begin{aligned} & 150 \\ & 0 \end{aligned}$ | $\begin{aligned} & 322 \\ & 3 \end{aligned}$ | $\begin{aligned} & 100 \\ & 92 \end{aligned}$ | $\begin{aligned} & 843 \\ & 3 \end{aligned}$ | $\begin{aligned} & 596 \\ & 5 \end{aligned}$ | $\begin{aligned} & 698 \\ & 9 \end{aligned}$ | $\begin{aligned} & 616 \\ & 9 \end{aligned}$ | $\begin{aligned} & 209 \\ & 5 \end{aligned}$ | $\begin{aligned} & 351 \\ & 9 \end{aligned}$ | $\begin{aligned} & 233 \\ & 3 \end{aligned}$ | $\begin{aligned} & 183 \\ & 5 \end{aligned}$ | $\begin{aligned} & 10 \\ & 14 \end{aligned}$ | $\begin{aligned} & 168 \\ & 3 \end{aligned}$ | $\begin{aligned} & 26 \\ & 7 \end{aligned}$ | $\begin{aligned} & 116 \\ & 5 \end{aligned}$ | $\begin{aligned} & 10 \\ & 14 \end{aligned}$ | $\begin{aligned} & 135 \\ & 2 \end{aligned}$ | ${ }_{2}^{121}$ | $\begin{aligned} & 56 \\ & 8 \end{aligned}$ | $\begin{aligned} & 78 \\ & 0 \end{aligned}$ | $\begin{aligned} & 87 \\ & 3 \end{aligned}$ | $\begin{aligned} & 44 \\ & 1 \end{aligned}$ | $\begin{aligned} & 24 \\ & 5 \end{aligned}$ | $\begin{aligned} & 22 \\ & 5 \end{aligned}$ | $\begin{aligned} & 48 \\ & 8 \end{aligned}$ | 55 2 |
| IGFS | $\begin{aligned} & 20 \\ & 12 \end{aligned}$ | 64 | 302 | 673 | 754 | $\begin{aligned} & 177 \\ & 4 \end{aligned}$ | $\begin{aligned} & 219 \\ & 7 \end{aligned}$ | $\begin{aligned} & { }_{1}^{720} \\ & 1 \end{aligned}$ | $\begin{aligned} & 842 \\ & 1 \end{aligned}$ | $\begin{aligned} & 710 \\ & 4 \end{aligned}$ | $\begin{aligned} & 102 \\ & 72 \end{aligned}$ | $\begin{aligned} & 947 \\ & 6 \end{aligned}$ | $\begin{aligned} & 313 \\ & 4 \end{aligned}$ | $\begin{aligned} & 674 \\ & 1 \end{aligned}$ | $\begin{aligned} & 397 \\ & 2 \end{aligned}$ | $\begin{aligned} & 383 \\ & 4 \end{aligned}$ | $\begin{aligned} & 17 \\ & 36 \end{aligned}$ | $\begin{aligned} & 290 \\ & 7 \end{aligned}$ | $\begin{aligned} & 54 \\ & 8 \end{aligned}$ | $\begin{aligned} & 236 \\ & 0 \end{aligned}$ | $\begin{aligned} & 17 \\ & 36 \end{aligned}$ | $\begin{aligned} & 244 \\ & 7 \end{aligned}$ | $\begin{aligned} & 251 \\ & 8 \end{aligned}$ | $\begin{aligned} & 10 \\ & 96 \end{aligned}$ | $\begin{aligned} & 14 \\ & 91 \end{aligned}$ | $\begin{aligned} & 18 \\ & 07 \end{aligned}$ | $\begin{aligned} & 78 \\ & 1 \end{aligned}$ | $\begin{aligned} & 49 \\ & 8 \end{aligned}$ | $\begin{aligned} & 39 \\ & 2 \end{aligned}$ | $\begin{aligned} & 99 \\ & 1 \end{aligned}$ | 85 0 |
| IGFS | $\begin{aligned} & 20 \\ & 13 \end{aligned}$ | 21 | 373 | 862 | $\begin{aligned} & 124 \\ & 3 \end{aligned}$ | $\begin{aligned} & 302 \\ & 6 \end{aligned}$ | $\begin{aligned} & 390 \\ & 3 \end{aligned}$ | $\begin{aligned} & 109 \\ & 18 \end{aligned}$ | $\begin{aligned} & 132 \\ & 84 \end{aligned}$ | $\begin{aligned} & 106 \\ & 90 \end{aligned}$ | $\begin{aligned} & 189 \\ & 29 \end{aligned}$ | $\begin{aligned} & 175 \\ & 31 \end{aligned}$ | $\begin{aligned} & 548 \\ & 3 \end{aligned}$ | $\begin{aligned} & 136 \\ & 36 \end{aligned}$ | $\begin{aligned} & 717 \\ & 7 \end{aligned}$ | $\begin{aligned} & 847 \\ & 1 \end{aligned}$ | $\begin{aligned} & 28 \\ & 78 \end{aligned}$ | $\begin{aligned} & 516 \\ & 5 \end{aligned}$ | $\begin{aligned} & 98 \\ & 0 \end{aligned}$ | $\begin{aligned} & 494 \\ & 1 \end{aligned}$ | $\begin{aligned} & 28 \\ & 78 \end{aligned}$ | $\begin{aligned} & 453 \\ & 0 \end{aligned}$ | $\begin{aligned} & 526 \\ & 5 \end{aligned}$ | $\begin{aligned} & 17 \\ & 84 \end{aligned}$ | $\begin{aligned} & 29 \\ & 64 \end{aligned}$ | $\begin{aligned} & 36 \\ & 13 \end{aligned}$ | $\begin{aligned} & 13 \\ & 12 \end{aligned}$ | $\begin{aligned} & 94 \\ & 1 \end{aligned}$ | $\begin{aligned} & 66 \\ & 6 \end{aligned}$ | $\begin{aligned} & 18 \\ & 62 \end{aligned}$ | $\begin{aligned} & 12 \\ & 91 \end{aligned}$ |
| IGFS | $\begin{aligned} & 20 \\ & 14 \end{aligned}$ | $\begin{aligned} & 13 \\ & 2 \end{aligned}$ | 28 | 47 | 90 | 423 | 794 | $\begin{aligned} & 295 \\ & 8 \end{aligned}$ | $\begin{aligned} & 442 \\ & 9 \end{aligned}$ | $\begin{aligned} & 369 \\ & 7 \end{aligned}$ | $\begin{aligned} & 745 \\ & 0 \end{aligned}$ | $\begin{aligned} & 712 \\ & 7 \end{aligned}$ | $\begin{aligned} & 221 \\ & 3 \end{aligned}$ | $\begin{aligned} & 596 \\ & 5 \end{aligned}$ | $\begin{aligned} & 287 \\ & 3 \end{aligned}$ | $\begin{aligned} & 381 \\ & 8 \end{aligned}$ | $\begin{aligned} & 12 \\ & 48 \end{aligned}$ | $\begin{aligned} & 214 \\ & 6 \end{aligned}$ | $\begin{aligned} & 49 \\ & 9 \end{aligned}$ | $\begin{aligned} & 223 \\ & 6 \end{aligned}$ | $\begin{aligned} & 12 \\ & 48 \end{aligned}$ | $\begin{aligned} & 196 \\ & 7 \end{aligned}$ | $\begin{aligned} & 243 \\ & 7 \end{aligned}$ | $\begin{aligned} & 88 \\ & 3 \end{aligned}$ | $\begin{aligned} & 13 \\ & 17 \end{aligned}$ | $\begin{aligned} & 17 \\ & 17 \end{aligned}$ | 59 8 | $\begin{aligned} & 48 \\ & 0 \end{aligned}$ | $\begin{aligned} & 30 \\ & 8 \end{aligned}$ | $\begin{aligned} & 94 \\ & 1 \end{aligned}$ | 47 8 |
| IGFS | $\begin{aligned} & 20 \\ & 15 \end{aligned}$ | 30 | 815 | $347$ | $\begin{aligned} & 137 \\ & 7 \end{aligned}$ | 516 | 943 | $\begin{aligned} & 484 \\ & 5 \end{aligned}$ | $\begin{aligned} & 745 \\ & 4 \end{aligned}$ | $\begin{aligned} & 585 \\ & 8 \end{aligned}$ | $\begin{aligned} & 140 \\ & 16 \end{aligned}$ | $\begin{aligned} & 166 \\ & 39 \end{aligned}$ | $\begin{aligned} & 462 \\ & 3 \end{aligned}$ | $\begin{aligned} & 135 \\ & 24 \end{aligned}$ | $\begin{aligned} & 524 \\ & 3 \end{aligned}$ | $\begin{aligned} & 903 \\ & 0 \end{aligned}$ | $\begin{aligned} & 39 \\ & 79 \end{aligned}$ | $\begin{aligned} & 449 \\ & 4 \end{aligned}$ | $\begin{aligned} & 16 \\ & 90 \end{aligned}$ | $\begin{aligned} & 643 \\ & 8 \end{aligned}$ | $\begin{aligned} & 39 \\ & 79 \end{aligned}$ | $548$ | $\begin{aligned} & 639 \\ & 3 \end{aligned}$ | 39 90 | $\begin{aligned} & 49 \\ & 77 \end{aligned}$ | $\begin{aligned} & 48 \\ & 86 \end{aligned}$ | 34 70 | $\begin{aligned} & 17 \\ & 67 \end{aligned}$ | $\begin{aligned} & 20 \\ & 01 \end{aligned}$ | $\begin{aligned} & 30 \\ & 02 \end{aligned}$ | 74 3 |
| IGFS | $\begin{aligned} & 20 \\ & 16 \end{aligned}$ | $\begin{aligned} & 21 \\ & 5 \end{aligned}$ | 282 | $\begin{aligned} & 240 \\ & 0 \end{aligned}$ | $\begin{aligned} & 288 \\ & 8 \end{aligned}$ | $\begin{aligned} & 268 \\ & 2 \end{aligned}$ | $\begin{aligned} & 176 \\ & 1 \end{aligned}$ | $\begin{aligned} & 445 \\ & 8 \end{aligned}$ | $\begin{aligned} & 777 \\ & 3 \end{aligned}$ | $\begin{aligned} & 617 \\ & 3 \end{aligned}$ | $\begin{aligned} & 160 \\ & 77 \end{aligned}$ | $\begin{aligned} & 177 \\ & 88 \end{aligned}$ | $\begin{aligned} & 538 \\ & 6 \end{aligned}$ | $\begin{aligned} & 162 \\ & 40 \end{aligned}$ | $\begin{aligned} & 606 \\ & 6 \end{aligned}$ | $\begin{aligned} & 109 \\ & 38 \end{aligned}$ | $\begin{aligned} & 42 \\ & 31 \end{aligned}$ | $\begin{aligned} & 530 \\ & 2 \end{aligned}$ | $\begin{aligned} & 22 \\ & 26 \end{aligned}$ | $\begin{aligned} & 738 \\ & 9 \end{aligned}$ | $\begin{aligned} & 42 \\ & 31 \end{aligned}$ | $\begin{aligned} & 603 \\ & 6 \end{aligned}$ | $\begin{aligned} & 806 \\ & 2 \end{aligned}$ | 48 80 | $\begin{aligned} & 49 \\ & 10 \end{aligned}$ | $\begin{aligned} & 62 \\ & 58 \end{aligned}$ | 31 05 | $\begin{aligned} & 19 \\ & 02 \end{aligned}$ | $\begin{aligned} & 15 \\ & 95 \end{aligned}$ | $\begin{aligned} & 37 \\ & 19 \end{aligned}$ | 81 9 |
| IGFS | $\begin{aligned} & 20 \\ & 17 \end{aligned}$ | $\begin{aligned} & 10 \\ & 22 \\ & 8 \end{aligned}$ | $\begin{aligned} & 696 \\ & 697 \end{aligned}$ | $\begin{aligned} & 608 \\ & 0 \end{aligned}$ | $\begin{aligned} & 932 \\ & 2 \end{aligned}$ | $\begin{aligned} & 164 \\ & 17 \end{aligned}$ | $\begin{aligned} & 113 \\ & 47 \end{aligned}$ | $\begin{aligned} & 958 \\ & 5 \end{aligned}$ | $\begin{aligned} & 881 \\ & 81 \end{aligned}$ | $\begin{aligned} & 585 \\ & 3 \end{aligned}$ | $\begin{aligned} & 127 \\ & 38 \end{aligned}$ | $\begin{aligned} & 137 \\ & 21 \end{aligned}$ | $\begin{aligned} & 443 \\ & 6 \end{aligned}$ | $\begin{aligned} & 126 \\ & 70 \end{aligned}$ | $\begin{aligned} & 456 \\ & 4 \end{aligned}$ | $\begin{aligned} & 847 \\ & 5 \end{aligned}$ | $\begin{aligned} & 39 \\ & 44 \end{aligned}$ | $\begin{aligned} & 419 \\ & 5 \end{aligned}$ | $\begin{aligned} & 19 \\ & 23 \end{aligned}$ | $\begin{aligned} & 627 \\ & 8 \end{aligned}$ | $\begin{aligned} & 39 \\ & 44 \end{aligned}$ | $\begin{aligned} & 526 \\ & 6 \end{aligned}$ | $\begin{aligned} & 649 \\ & 0 \end{aligned}$ | 46 24 | $\begin{aligned} & 47 \\ & 44 \end{aligned}$ | $\begin{aligned} & 51 \\ & 68 \end{aligned}$ | 34 22 | $\begin{aligned} & 17 \\ & 78 \end{aligned}$ | $\begin{aligned} & 18 \\ & 96 \end{aligned}$ | $\begin{aligned} & 31 \\ & 86 \end{aligned}$ | 64 0 |
| $\begin{gathered} \text { SPN } \\ \text { GFS } \end{gathered}$ | $\begin{aligned} & 19 \\ & 91 \end{aligned}$ | 1 | $\begin{aligned} & 140 \\ & 2 \end{aligned}$ | 881 | 102 | 15 | 6 | 5 | 3 | 2 | 2 | 2 | $\begin{aligned} & 0.6 \\ & 2 \end{aligned}$ | $\begin{aligned} & -0.9 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0.7 \\ & 8 \end{aligned}$ | 0.5 | $\begin{aligned} & 0 . \\ & 18 \end{aligned}$ | $\begin{aligned} & 0.4 \\ & 8 \end{aligned}$ |  | $\begin{aligned} & 0.2 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 18 \end{aligned}$ | 0.3 | $\begin{aligned} & 0.2 \\ & 5 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 12 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 12 \end{aligned}$ |  | 3 | 3 |  | $\begin{aligned} & 0 . \\ & 18 \end{aligned}$ |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | 19 92 | $\begin{aligned} & 10 \\ & 4 \end{aligned}$ | $\begin{aligned} & 460 \\ & 9 \end{aligned}$ | $\begin{aligned} & 183 \\ & 0 \end{aligned}$ | 95 | 17 | 13 | 41 | 53 | 35 | 103 | 156 | 57 | 175 | 37 | 120 | 64 | 56 | 45 | 94 | 64 | 76 | 114 | 98 | 61 | $\begin{aligned} & 10 \\ & 2 \end{aligned}$ | 49 | 35 | 25 | 71 | 4 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 19 \\ & 93 \end{aligned}$ | $\begin{aligned} & 17 \\ & 51 \end{aligned}$ | $\begin{aligned} & 550 \\ & 8 \end{aligned}$ | $\begin{aligned} & 242 \\ & 4 \end{aligned}$ | 163 | 49 | 18 | 5 | 3 | 2 | 2 | 2 | $\begin{aligned} & 0.6 \\ & 4 \end{aligned}$ | 1 | $\begin{aligned} & 0.7 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.5 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 28 \end{aligned}$ | $\begin{aligned} & 0.5 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 09 \end{aligned}$ | $\begin{aligned} & 0.2 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 28 \end{aligned}$ | $\begin{aligned} & 0.3 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.3 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 0 . \end{aligned}$ | $\begin{aligned} & 0 . \\ & 09 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 28 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 09 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 18 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 19 \end{aligned}$ |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 19 \\ & 94 \end{aligned}$ | 73 | $\begin{aligned} & 105 \\ & 76 \end{aligned}$ | $\begin{aligned} & 124 \\ & 11 \end{aligned}$ | $\begin{aligned} & 384 \\ & 5 \end{aligned}$ | 643 | 57 | 35 | 17 | 5 | 5 | 4 | 1 | 3 | 1 | 2 | $\begin{aligned} & 0 . \\ & 27 \end{aligned}$ | $\begin{aligned} & 0.8 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 05 \end{aligned}$ | $\begin{aligned} & 0.8 \\ & 2 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 27 \end{aligned}$ | $\begin{aligned} & 0.6 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.8 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 05 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 39 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 48 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 05 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 09 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 22 \end{aligned}$ |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 19 \\ & 95 \end{aligned}$ | $\begin{aligned} & 19 \\ & 6 \end{aligned}$ | $\begin{aligned} & 423 \\ & 1 \end{aligned}$ | $\begin{aligned} & 152 \\ & 6 \end{aligned}$ | 107 | 66 | 51 | 64 | 48 | 30 | 41 | 35 | 11 | 22 | 13 | 13 | 4 | 9 | $\begin{aligned} & 0 . \\ & 9 \end{aligned}$ | 7 | 4 | 7 | 7 | 1 | 4 | 5 | $\begin{aligned} & 0 . \\ & 83 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 41 \end{aligned}$ | 2 | 3 |
| SPN | 19 | 89 | 670 | 290 | 584 | 553 | 254 | 109 | 66 | 38 | 72 | 67 | 20 | 53 | 23 | 36 | 11 | 17 | 5 | 22 | 11 | 18 | 23 | 9 | 15 | 16 | 8 | 4 | 4 | 9 | 3 |


| Sur vey - | $\begin{aligned} & \text { Ye } \\ & \text { ar } \end{aligned}$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GFS | 96 | 7 | 7 | 8 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 19 \\ & 97 \end{aligned}$ | $\begin{aligned} & 13 \\ & 51 \end{aligned}$ | $\begin{aligned} & 730 \\ & 6 \end{aligned}$ | $544$ | $\begin{aligned} & 160 \\ & 9 \end{aligned}$ | 681 | 249 | 203 | 121 | 67 | 69 | 56 | 18 | 22 | 18 | 11 | 4 | 11 | $\begin{aligned} & 0 . \\ & 14 \end{aligned}$ | 6 | 4 | 7 | 6 | $\begin{aligned} & 0 . \\ & 14 \end{aligned}$ | 3 | 3 |  | $\begin{aligned} & 0 . \\ & 14 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 27 \end{aligned}$ | 4 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 19 \\ & 98 \end{aligned}$ | 13 | $\begin{aligned} & 449 \\ & 3 \end{aligned}$ | $\begin{aligned} & 364 \\ & 0 \end{aligned}$ | 638 | 175 | 101 | 79 | 58 | 37 | 54 | 53 | 17 | 40 | 19 | 25 | 9 | 15 | 4 | 14 | 9 | 13 | 17 | 6 | 7 | 12 | 3 | 5 | 3 | 8 | 4 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 19 \\ & 99 \end{aligned}$ | 79 | $\begin{aligned} & 425 \\ & 8 \end{aligned}$ | ${ }_{2}^{180}$ | 116 | 93 | 80 | 112 | 121 | 85 | 191 | 195 | 61 | 175 | 70 | 117 | 35 | 58 | 18 | 65 | 35 | 55 | 77 | 25 | 34 | 57 | 14 | 18 | 7 | 37 | 10 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 00 \end{aligned}$ | $\begin{aligned} & 57 \\ & 82 \end{aligned}$ | $\begin{aligned} & 166 \\ & 1 \end{aligned}$ | $\begin{aligned} & 132 \\ & 5 \end{aligned}$ | 347 | 518 | 553 | 750 | 537 | 315 | 443 | 379 | 116 | 237 | 139 | 146 | 37 | 91 | 10 | 78 | 37 | 69 | 85 | 18 | 39 | 53 | 7 | 9 | 3 | 18 | 25 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 01 \end{aligned}$ | 73 | $\begin{aligned} & 595 \\ & 2 \end{aligned}$ | $\begin{aligned} & 309 \\ & 9 \end{aligned}$ | 308 | 205 | 161 | 197 | 190 | 148 | 199 | 175 | 58 | 114 | 77 | 62 | 25 | 53 | 6 | 34 | 25 | 38 | 38 | 11 | 17 | 25 | 4 | 5 | 2 | 11 | 17 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | 20 02 | 24 | $\begin{aligned} & 331 \\ & 5 \end{aligned}$ | $\begin{aligned} & 139 \\ & 5 \end{aligned}$ | 104 | 54 | 43 | 55 | 63 | 47 | 98 | 88 | 26 | 71 | 37 | 46 | 10 | 25 | 3 | 24 | 10 | 20 | 26 | 4 | 12 | 16 | 2 | 3 | $\begin{aligned} & 0 . \\ & 91 \end{aligned}$ | 7 | 6 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 03 \end{aligned}$ | $\begin{aligned} & 15 \\ & 21 \end{aligned}$ | 203 | 155 | 38 | 26 | 16 | 14 | 10 | 5 | 9 | 9 | 3 | 7 | 3 | 4 | 2 | 2 | $\begin{aligned} & 0 . \\ & 83 \end{aligned}$ | 3 | 2 | 2 | 3 | 2 | 2 | 2 | 1 | $\begin{aligned} & 0 . \\ & 73 \end{aligned}$ | $\begin{gathered} 0 . \\ 5 \end{gathered}$ | 1 | $\begin{aligned} & 0 . \\ & 42 \end{aligned}$ |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 04 \end{aligned}$ | 32 | $\begin{aligned} & 426 \\ & 7 \end{aligned}$ | $\begin{aligned} & 224 \\ & 3 \end{aligned}$ | 177 | 82 | 68 | 171 | 219 | 186 | 303 | 279 | 89 | 209 | 118 | 124 | 37 | 85 | 14 | 63 | 37 | 61 | 76 | 14 | 25 | 52 | $\begin{aligned} & 0 . \\ & 4 \end{aligned}$ | 14 | $\begin{aligned} & 0 . \\ & 2 \end{aligned}$ | 28 | 23 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 05 \end{aligned}$ | $\begin{aligned} & 49 \\ & 2 \end{aligned}$ | $\begin{aligned} & 125 \\ & 3 \end{aligned}$ | 701 | 108 | 78 | 46 | 50 | 60 | 51 | 84 | 78 | 25 | 59 | 33 | 35 | 15 | 24 | 4 | 22 | 15 | 22 | 22 | 9 | 16 | 15 | 9 | 4 | 4 | 8 | 6 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 06 \end{aligned}$ | $\begin{aligned} & 33 \\ & 0 \end{aligned}$ | ${ }^{729}$ | $\begin{aligned} & 737 \\ & 8 \end{aligned}$ | $119$ | 85 | 34 | 36 | 56 | 44 | 116 | 112 | 33 | 100 | 43 | 68 | 14 | 32 | 8 | 35 | 14 | 27 | 42 | 9 | 15 | 29 | 2 | 8 | $\begin{aligned} & 0 . \\ & 9 \end{aligned}$ | 15 | 6 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 07 \end{aligned}$ | 90 | $\begin{aligned} & 664 \\ & 6 \end{aligned}$ | $\begin{aligned} & 399 \\ & 0 \end{aligned}$ | 367 | 180 | 106 | 37 | 30 | 18 | 55 | 54 | 16 | 50 | 20 | 35 | 8 | 15 | 4 | 20 | 8 | 15 | 22 | 7 | 11 | 15 | 4 | 4 | 2 | 8 | 2 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 08 \end{aligned}$ | $\begin{aligned} & 34 \\ & 3 \end{aligned}$ | ${ }_{6}^{173}$ | $\begin{aligned} & 188 \\ & 6 \end{aligned}$ | 629 | 908 | 597 | 329 | 178 | 62 | 202 | 183 | 47 | 158 | 53 | 122 | 28 | 36 | 10 | 81 | 28 | 54 | 73 | 32 | 63 | 47 | 37 | 9 | 19 | 18 | $\begin{aligned} & 0 . \\ & 28 \end{aligned}$ |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 09 \end{aligned}$ | $\begin{aligned} & 19 \\ & 5 \end{aligned}$ | $\begin{aligned} & 448 \\ & 7 \end{aligned}$ | $\begin{aligned} & 507 \\ & 7 \end{aligned}$ | $\begin{aligned} & 108 \\ & 5 \end{aligned}$ | 168 | 104 | 79 | 71 | 26 | 174 | 155 | 37 | 147 | 56 | 113 | 9 | 34 | 6 | 58 | 9 | 34 | 62 | 8 | 29 | 37 | 3 | 6 | 2 | 11 | 1 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 10 \end{aligned}$ | $\begin{aligned} & 16 \\ & 2 \end{aligned}$ | $\begin{aligned} & 245 \\ & 58 \end{aligned}$ | $\begin{aligned} & 135 \\ & 72 \end{aligned}$ | $\begin{aligned} & 150 \\ & 4 \end{aligned}$ | 792 | 346 | 101 | 85 | 41 | 222 | 365 | 132 | 436 | 76 | 306 | $\begin{aligned} & 14 \\ & 6 \end{aligned}$ | 130 | 91 | 206 | $\begin{aligned} & 14 \\ & 6 \end{aligned}$ | 178 | 245 | $\begin{aligned} & 14 \\ & 6 \end{aligned}$ | $\begin{aligned} & 13 \\ & 5 \end{aligned}$ | $\begin{aligned} & 21 \\ & 3 \end{aligned}$ | $\begin{aligned} & 10 \\ & 4 \end{aligned}$ | 90 | 52 | $\begin{aligned} & 18 \\ & 0 \end{aligned}$ | 4 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 11 \end{aligned}$ | $\begin{aligned} & 39 \\ & 3 \end{aligned}$ | $\begin{aligned} & 573 \\ & 0 \end{aligned}$ | $\begin{aligned} & 365 \\ & 6 \end{aligned}$ | 432 | 244 | 163 | 94 | 77 | 38 | 140 | 182 | 61 | 198 | 48 | 140 | 50 | 59 | 33 | 84 | 50 | 68 | 103 | 48 | 45 | 85 | 27 | 33 | 14 | 66 | 4 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 12 \end{aligned}$ | $\begin{aligned} & 19 \\ & 6 \end{aligned}$ | $\begin{aligned} & 116 \\ & 53 \end{aligned}$ | $\begin{aligned} & 535 \\ & 9 \end{aligned}$ | 383 | 62 | 55 | 160 | 276 | 202 | 620 | 657 | 201 | 638 | 228 | 441 | $\begin{aligned} & 14 \\ & 0 \end{aligned}$ | 198 | 73 | 266 | $\begin{aligned} & 14 \\ & 0 \end{aligned}$ | 215 | 295 | $\begin{aligned} & 12 \\ & 2 \end{aligned}$ | $\begin{aligned} & 16 \\ & 1 \end{aligned}$ | $\begin{aligned} & 22 \\ & 0 \end{aligned}$ | 86 | 71 | 43 | $\begin{aligned} & 14 \\ & 1 \end{aligned}$ | 26 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 13 \end{aligned}$ | 13 | $\begin{aligned} & 476 \\ & 3 \end{aligned}$ | $\begin{aligned} & 294 \\ & 7 \end{aligned}$ | 446 | 439 | 276 | 110 | 59 | 30 | 44 | 49 | 17 | 44 | 16 | 28 | 15 | 16 | 7 | 21 | 15 | 19 | 22 | 16 | 17 | 18 | 13 | 6 | 6 | 13 | 3 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 14 \end{aligned}$ | $\begin{aligned} & 19 \\ & 8 \end{aligned}$ | 542 | 611 | 767 | $\begin{aligned} & 113 \\ & 1 \end{aligned}$ | 910 | 875 | 626 | 323 | 711 | 914 | 317 | 926 | 228 | 635 | $\begin{aligned} & 27 \\ & 1 \end{aligned}$ | 291 | $\begin{aligned} & 16 \\ & 8 \end{aligned}$ | 402 | $\begin{aligned} & 27 \\ & 1 \end{aligned}$ | 348 | 488 | $\begin{aligned} & 25 \\ & 9 \end{aligned}$ | $\begin{aligned} & 24 \\ & 0 \end{aligned}$ | $\begin{aligned} & 41 \\ & 2 \end{aligned}$ | $\begin{aligned} & 16 \\ & 3 \end{aligned}$ | $\begin{aligned} & 16 \\ & 5 \end{aligned}$ | 82 | $\begin{aligned} & 32 \\ & 9 \end{aligned}$ | 25 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 15 \end{aligned}$ | 83 | $\begin{aligned} & 420 \\ & 7 \end{aligned}$ | $\begin{aligned} & 243 \\ & 0 \end{aligned}$ | 248 | 463 | 516 | 616 | 432 | 233 | 403 | 463 | 158 | 419 | 125 | 281 | $\begin{aligned} & 13 \\ & 0 \end{aligned}$ | 138 | 74 | 193 | $\begin{aligned} & 13 \\ & 0 \end{aligned}$ | 166 | 221 | $\begin{aligned} & 14 \\ & 0 \end{aligned}$ | $\begin{aligned} & 12 \\ & 7 \end{aligned}$ | $\begin{aligned} & 18 \\ & 5 \end{aligned}$ | 91 | 67 | 46 | $\begin{aligned} & 13 \\ & 4 \end{aligned}$ | 17 |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 16 \end{aligned}$ | 1 | 23 | 17 | 7 | 7 | 4 | 4 | 2 | 1 | 2 | 2 | $\begin{aligned} & 0.5 \\ & 9 \end{aligned}$ | 1 | $\begin{aligned} & 0.6 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.8 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 21 \end{aligned}$ | $\begin{aligned} & 0.4 \\ & 8 \end{aligned}$ | 0. 08 | $\begin{aligned} & 0.4 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 21 \end{aligned}$ | $\begin{aligned} & 0.3 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.5 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 11 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 23 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 33 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 05 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 07 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 03 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 15 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 11 \end{aligned}$ |
| $\begin{aligned} & \text { SPN } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 17 \end{aligned}$ | $\begin{aligned} & 0.3 \\ & 8 \end{aligned}$ | 16 | 14 | 3 | 2 | 2 | 3 | 2 | 1 | 3 | 3 | 1 | 3 | 1 | 2 | $\begin{aligned} & 0 . \\ & 69 \end{aligned}$ | $\begin{aligned} & 0.9 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 4 \end{aligned}$ | 1 | $\begin{aligned} & 0 . \\ & 69 \end{aligned}$ | $\begin{aligned} & 0.9 \\ & 9 \end{aligned}$ | 1 | $\begin{aligned} & 0 . \\ & 59 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 61 \end{aligned}$ | 1 | $\begin{aligned} & 0 . \\ & 31 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 41 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 17 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 79 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 14 \end{aligned}$ |
| SPP | 20 | 4 | 6 | 73 | 47 | 128 | 163 | 290 | 369 | 271 | 650 | 581 | 165 | 482 | 241 | 324 | 62 | 158 | 21 | 170 | 62 | 133 | 183 | 29 | 87 | 11 | 16 | 21 | 8 | 42 | 33 |


| $\begin{aligned} & \text { Sur } \\ & \text { vey } \end{aligned}$ | $\begin{aligned} & \mathrm{Ye} \\ & \text { ar } \end{aligned}$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GFS | 01 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 |  |  |  |  |  |
| SPP | 20 |  | 0.03 | 0.39 | 4 | 29 | 57 | 162 | 201 | 161 | 294 | 272 | 84 | 214 | 112 | 134 | 40 | 80 | 14 | 73 | 40 | 66 | 81 | 20 | 38 | 55 | 12 | 14 | 6 | 28 | 20 |
| GFS | 02 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 |  | 1 | 7 | 12 | 21 | 21 | 50 | 69 | 54 | 125 | 126 | 39 | 114 | 47 | 76 | 23 | 38 | 12 | 43 | 23 | 36 | 50 | 17 | 23 | 36 | 10 | 12 | 6 | 23 | 7 |
| GFS | 03 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 | 1 | 6 | 3 | 3 | 10 | 19 | 66 | 86 | 65 | 145 | 150 | 47 | 135 | 54 | 89 | 27 | 45 | 15 | 49 | 27 | 42 | 59 | 19 | 24 | 44 | 9 | 15 | 4 | 29 | 8 |
| GFS | 04 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 | 2 | 18 | 18 | 9 | 13 | 17 | 81 | 132 | 103 | 263 | 283 | 90 | 269 | 98 | 181 | 68 | 88 | 34 | 115 | 68 | 97 | 126 | 62 | 74 | 97 | 45 | 32 | 23 | 64 | 13 |
| GFS | 05 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 | 2 | 137 | 77 | 33 | 53 | 36 | 51 | 84 | 64 | 180 | 200 | 64 | 197 | 67 | 134 | 53 | 63 | 26 | 88 | 53 | 74 | 94 | 49 | 60 | 73 | 39 | 26 | 20 | 50 | 8 |
| GFS | 06 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 |  | 12 | 19 | 12 | 14 | 15 | 22 | 24 | 16 | 41 | 47 | 15 | 47 | 15 | 32 | 11 | 15 | 7 | 19 | 11 | 16 | 23 | 11 | 10 | 19 | 5 | 7 | 3 | 13 | 2 |
| GFS | 07 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 | 1 | 9 | 15 | 13 | 25 | 35 | 72 | 79 | 53 | 130 | 135 | 42 | 125 | 46 | 85 | 27 | 40 | 14 | 51 | 27 | 42 | 57 | 23 | 30 | 43 | 16 | 14 | 8 | 27 | 6 |
| GFS | 08 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 | 11 | 13 | 5 | 5 | 45 | 91 | 228 | 263 | 197 | 390 | 429 | 143 | 394 | 144 | 257 | 10 | 137 | 54 | 161 | 10 | 146 | 183 | 88 | 10 | 14 | 65 | 53 | 32 | 10 | 23 |
| GFS | 09 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 9 |  |  |  | 9 |  |  |  | 2 | 5 |  |  |  | 7 |  |
| SPP | 20 | 1 | 18 | 5 | 4 | 15 | 41 | 156 | 167 | 121 | 236 | 236 | 75 | 201 | 84 | 131 | 46 | 69 | 22 | 79 | 46 | 69 | 89 | 37 | 47 | 66 | 25 | 21 | 12 | 42 | 13 |
| GFS | 10 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 |  | 0.43 | 7 | 12 | 17 | 22 | 109 | 159 | 133 | 261 | 256 | 81 | 216 | 100 | 138 | 48 | 78 | 21 | 83 | 48 | 73 | 91 | 37 | 49 | 66 | 24 | 20 | 12 | 41 | 17 |
| GFS | 11 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 | 1 | 1 | 2 | 2 | 4 | 10 | 57 | 86 | 72 | 149 | 143 | 44 | 121 | 57 | 78 | 26 | 43 | 10 | 46 | 26 | 40 | 50 | 18 | 28 | 35 | 13 | 10 | 7 | 20 | 9 |
| GFS | 12 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 | 1 | 19 | 17 | 6 | 3 | 5 | 49 | 102 | 80 | 235 | 239 | 72 | 226 | 88 | 155 | 47 | 71 | 23 | 93 | 47 | 75 | 101 | 41 | 56 | 74 | 28 | 22 | 15 | 44 | 11 |
| GFS | 13 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 | 19 | 5 | 31 | 38 | 21 | 14 | 219 | 597 | 438 | 163 | 164 | 478 | 160 | 603 | 112 | 41 | 476 | 16 | 791 | 41 | 626 | 739 | 42 | 63 | 53 | 42 | 18 | 25 | 28 | 61 |
| GFS | 14 |  |  |  |  |  |  |  |  |  | 2 | 7 |  | 2 |  | 6 | 7 |  | 0 |  | 7 |  |  | 0 | 3 | 0 | 3 | 5 | 3 | 8 |  |
| SPP | 20 | 2 | 1 | 1 | 0.77 | 0.83 | 3 | 35 | 67 | 56 | 136 | 142 | 45 | 132 | 52 | 88 | 37 | 44 | 19 | 63 | 37 | 52 | 67 | 47 | 45 | 52 | 30 | 14 | 15 | 29 | 8 |
| GFS | 15 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPP | 20 | 0.0 | 0.02 | 0.05 | 0.09 | 0.06 | 0.0 | 0.1 | 0.4 | 0.3 | 1 | 1 | 0.3 | 1 | 0.4 | 0.7 | 0. | 0.3 | 0. | 0.5 | 0. | 0.4 | 0.5 | 0. | 0. | 0. | 0. | 0. | 0. | 0. | 0. |
| GFS | 16 | 4 |  |  |  |  | 3 | 9 | 6 | 6 |  |  | 6 |  | 2 | 9 | 28 | 6 | 15 | 3 | 28 | 2 | 7 | 34 | 35 | 44 | 22 | 13 | 11 | 25 | 05 |
| SPP | 20 | 2 | 0.12 | 0.08 | 0.01 | 0.11 | 0.1 | 0.5 | 0.8 | 0.5 | 2 | 3 | 0.9 | 3 | 0.8 | 2 | 1 | 0.9 | 0. | 2 | 1 | 1 | 2 | 1 | 1 | 1 | 1 | 0. | 0. | 0. | 0. |
| GFS | 17 |  |  |  |  |  | 9 |  | 9 | 7 |  |  | 3 |  | 3 |  |  | 2 | 5 |  |  |  |  |  |  |  |  | 47 | 52 | 94 | 07 |
| WCS | 19 |  |  | 0.38 | 0.12 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| GFS | 86 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WCS | 19 |  | 0.01 | 0.58 | 0.64 | 1 | 0.7 | 0.1 | 0.0 | 0.0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| GFS | 87 |  |  |  |  |  | 6 | 8 | 5 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WCS | 19 | 0.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| GFS | 88 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WCS | 19 |  | 0.3 | 0.2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| GFS | 89 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WCS | 19 | 1 | 2 | 10 | 21 | 46 | 39 | 31 | 16 | 7 | 5 | 4 | 1 | 0.7 | 0.9 | 0.1 | 0. | 0.6 |  | 0.0 | 0. | 0.3 | 0.0 |  | 0. | 0. |  |  |  |  | 0. |
| GFS | 90 |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 2 | 2 | 29 | 1 |  | 6 | 29 | 2 | 6 |  | 03 | 03 |  |  |  |  | 29 |
| WCS | 19 |  | 2 | 23 | 52 | 175 | 186 | 194 | 105 | 45 | 36 | 28 | 9 | 5 | 5 | 2 | 1 | 3 |  | 0.9 | 1 | 2 | 0.9 |  | 0. | 0. |  |  |  |  | 1 |


| Sur vey | $\begin{aligned} & \mathrm{Ye} \\ & \text { ar } \end{aligned}$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GFS | 91 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 7 |  |  | 7 |  | 48 | 48 |  |  |  |  |  |
| WCS GFS | $\begin{aligned} & 19 \\ & 92 \end{aligned}$ |  | 2 | 33 | 116 | 616 | 975 | $\begin{aligned} & 195 \\ & 2 \end{aligned}$ | $\begin{aligned} & 127 \\ & 0 \end{aligned}$ | 712 | 662 | 524 | 178 | 157 | 152 | 61 | 41 | 96 |  | 30 | 41 | 56 | 30 |  | 15 | 15 |  |  |  |  | 41 |
| WCS GFS | $\begin{aligned} & 19 \\ & 93 \end{aligned}$ |  | 2 | 3 | 4 | 23 | 41 | 80 | 52 | 29 | 26 | 21 | 7 | 6 | 6 | 2 | 2 | 4 |  | 1 | 2 | 2 | $\begin{aligned} & 0.9 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 0 . \end{aligned}$ | $\begin{aligned} & 0 . \\ & 58 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 48 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 05 \end{aligned}$ |  | 2 |
| WCS GFS | $\begin{aligned} & 19 \\ & 94 \end{aligned}$ |  | 0.01 | 0.15 | 0.34 | 0.48 | $\begin{aligned} & 0.3 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.1 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 2 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 0.0 \\ & 2 \end{aligned}$ |  | $\begin{aligned} & 0.0 \\ & 3 \end{aligned}$ |  | $\begin{aligned} & 0.0 \\ & 2 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 3 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 02 \end{aligned}$ | $\begin{aligned} & 0 . \\ & 02 \end{aligned}$ |  |  |  |  |  |
| WCS GFS | $\begin{aligned} & 19 \\ & 95 \end{aligned}$ |  | 0.2 | 3 | 15 | 74 | 113 | 189 | 151 | 103 | 121 | 101 | 33 | 54 | 42 | 27 | 11 | 27 | $\begin{aligned} & 0 . \\ & 98 \end{aligned}$ | 13 | 11 | 17 | 14 | $\begin{aligned} & 0 . \\ & 98 \end{aligned}$ | 6 | 8 |  | $\begin{aligned} & 0 . \\ & 98 \end{aligned}$ |  | 2 | 10 |
| WCS GFS | $\begin{aligned} & 19 \\ & 96 \end{aligned}$ | 2 | 5 | 2 | 0.03 | 1 | 6 | 67 | 153 | 112 | 391 | 353 | 95 | 318 | 144 | 224 | 29 | 93 | 14 | 112 | 29 | 78 | 126 | 14 | 49 | 77 |  | 14 |  | 28 | 15 |
| WCS GFS | $\begin{aligned} & 19 \\ & 97 \end{aligned}$ | 4 | 4 | 11 | 6 | 12 | 22 | 63 | 62 | 47 | 69 | 60 | 19 | 40 | 25 | 23 | 7 | 17 | 2 | 12 | 7 | 12 | 13 | 2 | 6 | 9 | $\begin{aligned} & 0 . \\ & 8 \end{aligned}$ | 2 | $\begin{aligned} & 0 . \\ & 4 \end{aligned}$ | 4 | 5 |
| WCS GFS | $\begin{aligned} & 19 \\ & 98 \end{aligned}$ | 1 | 4 | 4 | 0.67 | 1 | 1 |  | $\begin{aligned} & 0.6 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.5 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.4 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.3 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0.1 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.1 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.2 \\ & 2 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 08 \end{aligned}$ | $\begin{aligned} & 0.1 \\ & 5 \end{aligned}$ |  |  | $\begin{aligned} & 0 . \\ & 08 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 8 \end{aligned}$ |  |  |  |  |  |  |  |  | 0. 08 |
| WCS GFS | $\begin{aligned} & 19 \\ & 99 \end{aligned}$ | 1 | 5 | 3 | 0.8 | 0.47 | $\begin{aligned} & 0.5 \\ & 8 \end{aligned}$ | 1 | 0.7 | 0.4 | $\begin{aligned} & 0.3 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.2 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 8 \end{aligned}$ |  | $\begin{aligned} & 0 . \\ & 02 \end{aligned}$ | $\begin{aligned} & 0.0 \\ & 5 \end{aligned}$ |  |  | $\begin{aligned} & 0 . \\ & 02 \end{aligned}$ | ${ }_{2}^{0.0}$ |  |  |  |  |  |  |  |  | 0. |
| WCS GFS | $\begin{aligned} & 20 \\ & 00 \end{aligned}$ |  | 2 | 16 | 41 | 124 | 142 | 179 | 116 | 65 | 68 | 59 | 20 | 30 | 19 | 16 | 7 | 14 | 2 | 8 | 7 | 10 | 10 | 3 | 4 | 7 | 1 | 2 | $\begin{aligned} & 0 . \\ & 6 \end{aligned}$ | 4 | 5 |
| $\begin{aligned} & \text { WCS } \\ & \text { GFS } \end{aligned}$ | $\begin{aligned} & 20 \\ & 01 \end{aligned}$ | 1 | 0.11 | 2 | 5 | 17 | 21 | 40 | 44 | 30 | 70 | 67 | 20 | 58 | 25 | 39 | 9 | 19 | 5 | 21 | 9 | 17 | 25 | 7 | 10 | 18 | 2 | 5 | 1 | 9 | 3 |
| WCS GFS | $\begin{aligned} & 20 \\ & 02 \end{aligned}$ |  | 6 | 8 | 35 | 291 | 631 | $\begin{aligned} & 183 \\ & 8 \end{aligned}$ | $\begin{aligned} & 181 \\ & 4 \end{aligned}$ | $\begin{aligned} & 132 \\ & 0 \end{aligned}$ | $\begin{aligned} & 218 \\ & 4 \end{aligned}$ | $\begin{aligned} & 193 \\ & 5 \end{aligned}$ | 594 | $\begin{aligned} & 138 \\ & 6 \end{aligned}$ | 781 | 858 | $\begin{aligned} & 22 \\ & 5 \end{aligned}$ | 528 | 68 | 446 | $\begin{aligned} & 22 \\ & 5 \end{aligned}$ | 405 | 497 | 85 | $\begin{aligned} & 21 \\ & 4 \end{aligned}$ | $\begin{aligned} & 31 \\ & 7 \end{aligned}$ | 33 | 68 | 17 | $\begin{aligned} & 13 \\ & 6 \end{aligned}$ | 14 0 |
| WCS GFS | $\begin{aligned} & 20 \\ & 03 \end{aligned}$ | 1 | 2 | 42 | 28 | 127 | 272 | 867 | 971 | 691 | $\begin{aligned} & 149 \\ & 8 \end{aligned}$ | $\begin{aligned} & 151 \\ & 9 \end{aligned}$ | 476 | $\begin{aligned} & 133 \\ & 9 \end{aligned}$ | 536 | 892 | $\begin{aligned} & 24 \\ & 8 \end{aligned}$ | 446 | $\begin{aligned} & 14 \\ & 3 \end{aligned}$ | 480 | $\begin{aligned} & 24 \\ & 8 \end{aligned}$ | 401 | 592 | $\begin{aligned} & 18 \\ & 2 \end{aligned}$ | $\begin{aligned} & 21 \\ & 5 \end{aligned}$ | $\begin{aligned} & 43 \\ & 9 \end{aligned}$ | 62 | $\begin{aligned} & 14 \\ & 0 \end{aligned}$ | 31 | $\begin{aligned} & 28 \\ & 0 \end{aligned}$ | 77 |
| WCS GFS | $\begin{aligned} & 20 \\ & 04 \end{aligned}$ | 1 | 2 | 16 | 57 | 327 | 770 | $\begin{aligned} & 259 \\ & 0 \end{aligned}$ | $\begin{aligned} & 268 \\ & 6 \end{aligned}$ | $\begin{aligned} & 198 \\ & 3 \end{aligned}$ | $\begin{aligned} & 344 \\ & 7 \end{aligned}$ | $\begin{aligned} & 335 \\ & 9 \end{aligned}$ | $\begin{aligned} & 107 \\ & 9 \end{aligned}$ | $\begin{aligned} & 269 \\ & 3 \end{aligned}$ | $\begin{aligned} & 124 \\ & 0 \end{aligned}$ | $\begin{aligned} & 170 \\ & 7 \end{aligned}$ | $\begin{aligned} & 56 \\ & 9 \end{aligned}$ | 986 | $\begin{aligned} & 26 \\ & 7 \end{aligned}$ | 957 | $\begin{aligned} & 56 \\ & 9 \end{aligned}$ | 866 | $\begin{aligned} & 112 \\ & 9 \end{aligned}$ | $\begin{aligned} & 38 \\ & 7 \end{aligned}$ | $\begin{aligned} & 48 \\ & 7 \end{aligned}$ | $\begin{aligned} & 83 \\ & 2 \end{aligned}$ | $\begin{aligned} & 19 \\ & 0 \end{aligned}$ | $\begin{aligned} & 25 \\ & 9 \end{aligned}$ | 95 | $\begin{aligned} & 51 \\ & 8 \end{aligned}$ | 21 5 |
| WCS GFS | $\begin{aligned} & 20 \\ & 05 \end{aligned}$ | 2 | 15 | 19 | 19 | 53 | 93 | 276 | 325 | 236 | 519 | 501 | 153 | 429 | 188 | 286 | 76 | 144 | 37 | 156 | 76 | 130 | 180 | 51 | 79 | $\begin{aligned} & 12 \\ & 7 \end{aligned}$ | 26 | 36 | 13 | 72 | 27 |
| WCS GFS | $\begin{aligned} & 20 \\ & 06 \end{aligned}$ | 4 | 4 | 12 | 39 | 183 | 196 | 340 | 423 | 294 | 781 | 834 | 261 | 795 | 283 | 543 | $\begin{aligned} & 17 \\ & 2 \end{aligned}$ | 252 | $\begin{aligned} & 10 \\ & 0 \end{aligned}$ | 322 | $\begin{aligned} & 17 \\ & 2 \end{aligned}$ | 261 | 379 | $\begin{aligned} & 16 \\ & 5 \end{aligned}$ | $\begin{aligned} & 17 \\ & 6 \end{aligned}$ | $\begin{aligned} & 29 \\ & 0 \end{aligned}$ | 87 | 93 | 43 | $\begin{aligned} & 18 \\ & 6 \end{aligned}$ | 35 |
| WCS GFS | $\begin{aligned} & 20 \\ & 07 \end{aligned}$ | 4 | 3 | 14 | 56 | 339 | 638 | $\begin{aligned} & 170 \\ & 7 \end{aligned}$ | $\begin{aligned} & 172 \\ & 7 \end{aligned}$ | $\begin{aligned} & 122 \\ & 0 \end{aligned}$ | $\begin{aligned} & 230 \\ & 9 \end{aligned}$ | $\begin{aligned} & 238 \\ & 5 \end{aligned}$ | 775 | $\begin{aligned} & 205 \\ & 6 \end{aligned}$ | 820 | $\begin{aligned} & 134 \\ & 1 \end{aligned}$ | $\begin{aligned} & 52 \\ & 2 \end{aligned}$ | 715 | $\begin{aligned} & 25 \\ & 2 \end{aligned}$ | 835 | $\begin{aligned} & 52 \\ & 2 \end{aligned}$ | 738 | 934 | $\begin{aligned} & 43 \\ & 9 \end{aligned}$ | $\begin{aligned} & 52 \\ & 0 \end{aligned}$ | 71 9 | $\begin{aligned} & 30 \\ & 5 \end{aligned}$ | $\begin{aligned} & 24 \\ & 0 \end{aligned}$ | ${ }_{2}^{15}$ | $\begin{aligned} & 48 \\ & 0 \end{aligned}$ | 13 0 |
| WCS GFS | $\begin{aligned} & 20 \\ & 08 \end{aligned}$ | 2 | 41 | 110 | 208 | 689 | 989 | $\begin{aligned} & 232 \\ & 4 \end{aligned}$ | $\begin{aligned} & 305 \\ & 4 \end{aligned}$ | $\begin{aligned} & 208 \\ & 2 \end{aligned}$ | $\begin{aligned} & 601 \\ & 3 \end{aligned}$ | $\begin{aligned} & 666 \\ & 2 \end{aligned}$ | $\begin{aligned} & 210 \\ & { }_{8} \end{aligned}$ | $\begin{aligned} & 656 \\ & 0 \end{aligned}$ | ${ }_{4}^{216}$ | $\begin{aligned} & 451 \\ & 7 \end{aligned}$ | $\begin{aligned} & 17 \\ & 12 \end{aligned}$ | $\begin{aligned} & 204 \\ & 2 \end{aligned}$ | $\begin{aligned} & 89 \\ & 4 \end{aligned}$ | $\begin{aligned} & 294 \\ & 5 \end{aligned}$ | $\begin{aligned} & 17 \\ & 12 \end{aligned}$ | $\begin{aligned} & 242 \\ & 4 \end{aligned}$ | $\begin{aligned} & 321 \\ & 0 \end{aligned}$ | $\begin{aligned} & 16 \\ & 95 \end{aligned}$ | $\begin{aligned} & 19 \\ & 69 \end{aligned}$ | 24 99 | 12 58 | $\begin{aligned} & 87 \\ & 2 \end{aligned}$ | 66 4 | 16 73 | 24 7 |
| WCS GFS | $\begin{aligned} & 20 \\ & 09 \end{aligned}$ | 1 | 2 | 100 | 387 | $\begin{aligned} & 181 \\ & 7 \end{aligned}$ | $\begin{aligned} & 153 \\ & 8 \end{aligned}$ | 759 | 363 | 137 | 139 | 136 | 46 | 95 | 43 | 58 | 32 | 37 | 12 | 43 | 32 | 41 | 42 | 28 | 35 | 33 | 26 | 11 | 13 | 22 | 8 |
| WCS GFS | $\begin{aligned} & 20 \\ & 10 \end{aligned}$ |  |  |  | 17 | 160 | 347 | 785 | 626 | 398 | 580 | 549 | 179 | 394 | 189 | 245 | 87 | 149 | 41 | 140 | 87 | 130 | 166 | 64 | 72 | $\begin{aligned} & 12 \\ & 3 \end{aligned}$ | 30 | 38 | 15 | 75 | 35 |
| WCS GFS | $\begin{aligned} & 20 \\ & 11 \end{aligned}$ | 6 | 31 | 531 | $\begin{aligned} & 108 \\ & 6 \end{aligned}$ | $\begin{aligned} & 351 \\ & 4 \end{aligned}$ | $538$ | $\begin{aligned} & 102 \\ & 38 \end{aligned}$ | $\begin{aligned} & 736 \\ & 9 \end{aligned}$ | $\begin{aligned} & 458 \\ & 9 \end{aligned}$ | $\begin{aligned} & 492 \\ & 5 \end{aligned}$ | $\begin{aligned} & 415 \\ & 7 \end{aligned}$ | $\begin{aligned} & 140 \\ & 3 \end{aligned}$ | $\begin{aligned} & 200 \\ & 4 \end{aligned}$ | $\begin{aligned} & 148 \\ & 9 \end{aligned}$ | 988 | $\begin{aligned} & 47 \\ & 7 \end{aligned}$ | $\begin{aligned} & 101 \\ & 6 \end{aligned}$ | 93 | 520 | $\begin{aligned} & 47 \\ & 7 \end{aligned}$ | 678 | 590 | $\begin{aligned} & 12 \\ & 4 \end{aligned}$ | $\begin{aligned} & 24 \\ & 9 \end{aligned}$ | 38 <br> 8 | 47 | 91 | 24 | 18 2 | 36 2 |
| WCS GFS | $\begin{aligned} & 20 \\ & 12 \end{aligned}$ | 1 | 5 | 28 | 97 | 469 | $\begin{aligned} & 114 \\ & 8 \end{aligned}$ | $\begin{aligned} & 480 \\ & 4 \end{aligned}$ | ${ }_{2}^{646}$ | $\begin{aligned} & 529 \\ & 8 \end{aligned}$ | $\begin{aligned} & 999 \\ & 0 \end{aligned}$ | $\begin{aligned} & 107 \\ & 65 \end{aligned}$ | $\begin{aligned} & 361 \\ & 0 \end{aligned}$ | $\begin{aligned} & 963 \\ & 2 \end{aligned}$ | $\begin{aligned} & 381 \\ & 0 \end{aligned}$ | $\begin{aligned} & 615 \\ & 5 \end{aligned}$ | 34 87 | $\begin{aligned} & 347 \\ & 7 \end{aligned}$ | $\begin{aligned} & 13 \\ & 93 \end{aligned}$ | $\begin{aligned} & 481 \\ & 4 \end{aligned}$ | $\begin{aligned} & 34 \\ & 87 \end{aligned}$ | $\begin{aligned} & 440 \\ & 4 \end{aligned}$ | $\begin{aligned} & 462 \\ & 1 \end{aligned}$ | $\begin{aligned} & 34 \\ & 30 \end{aligned}$ | 40 89 | 37 03 | 31 71 | $\begin{aligned} & 14 \\ & 91 \end{aligned}$ | 18 34 | 24 85 | ${ }^{65}$ |
| WCS | 20 | 1 | 0.6 | 0.43 | 5 | 101 | 381 | 242 | 337 | 300 | 467 | 422 | 136 | 306 | 185 | 176 | 64 | 129 | 17 | 971 | 64 | 999 | 106 | 26 | 52 | 71 | 17 | 17 | 86 | 35 | 38 |



Table 3.6.3.1. Boarfish in ICES Subareas 27.6, 7, 8. Key parameter estimates from the exploratory Schaeffer state space surplus production model. Posterior parameter distributions are provided in Figure 3.6.3.5.

|  | MEAN | SD | $\mathbf{2 . 5}$ | $\mathbf{2 5}$ | $\mathbf{5 0}$ | $\mathbf{7 5}$ | $\mathbf{9 7 . 5}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| r | $3.69 \mathrm{e}-01$ | $1.91 \mathrm{e}-01$ | $4.84 \mathrm{e}-02$ | $2.28 \mathrm{e}-01$ | $3.55 \mathrm{e}-01$ | $4.89 \mathrm{e}-01$ | $7.84 \mathrm{e}-01$ |
| K | $6.62 \mathrm{e}+05$ | $4.17 \mathrm{e}+05$ | $3.09 \mathrm{e}+05$ | $4.45 \mathrm{e}+05$ | $5.50 \mathrm{e}+05$ | $7.17 \mathrm{e}+05$ | $1.78 \mathrm{e}+06$ |
| F $_{\text {MSY }}$ | $1.85 \mathrm{e}-01$ | $9.53 \mathrm{e}-02$ | $2.42 \mathrm{e}-02$ | $1.14 \mathrm{e}-01$ | $1.78 \mathrm{e}-01$ | $2.45 \mathrm{e}-01$ | $3.92 \mathrm{e}-01$ |
| BMSY | $1.65 \mathrm{e}+05$ | $1.04 \mathrm{e}+05$ | $7.72 \mathrm{e}+04$ | $1.11 \mathrm{e}+05$ | $1.38 \mathrm{e}+05$ | $1.79 \mathrm{e}+05$ | $4.44 \mathrm{e}+05$ |
| TSB | $3.10 \mathrm{e}+05$ | $1.86 \mathrm{e}+05$ | $1.47 \mathrm{e}+05$ | $2.17 \mathrm{e}+05$ | $2.72 \mathrm{e}+05$ | $3.49 \mathrm{e}+05$ | $6.56 \mathrm{e}+05$ |

Table 3.6.4.1. Boarfish in ICES Subareas 27.6, 7, 8. Pseudo-cohort derived estimates of fishing mortality ( $\mathbf{F}$ ) and total mortality ( Z ), in comparison with total catch per year. Pearson correlation coefficient of $F$ vs. catch (tonnes) indicated.

| Age | Raised numbers |  |  |  |  |  |  |  |  |  |  | Ln raised numbers |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| 1 | 0 | 0 | 1575 | 2415 | 0 | 28 | 301 | 0 | 5556 | 218 | 1862 | 0 | 0 | 7 | 8 | 0 | 3 | 6 | 0 | 9 | 5 | 8 |
| 2 | 352 | 5488 | 15043 | 11229 | 2894 | 893 | 7148 | 695 | 116135 | 2385 | 4387 | 6 | 9 | 10 | 9 | 8 | 7 | 9 | 7 | 12 | 8 | 8 |
| 3 | 2114 | 21140 | 65744 | 72709 | 41913 | 5467 | 156680 | 49503 | 32248 | 10737 | 8830 | 8 | 10 | 11 | 11 | 11 | 9 | 12 | 11 | 10 | 9 | 9 |
| 4 | 40851 | 105575 | 338931 | 294382 | 28148 | 41278 | 58522 | 127520 | 16588 | 25114 | 34448 | 11 | 12 | 13 | 13 | 10 | 11 | 11 | 12 | 10 | 10 | 10 |
| 5 | 48915 | 141300 | 475619 | 567689 | 30116 | 110272 | 59797 | 93705 | 24564 | 20263 | 27266 | 11 | 12 | 13 | 13 | 10 | 12 | 11 | 11 | 10 | 10 | 10 |
| 6 | 62713 | 195339 | 543707 | 878363 | 175696 | 146582 | 68949 | 67275 | 26566 | 18025 | 21103 | 11 | 12 | 13 | 14 | 12 | 12 | 11 | 11 | 10 | 10 | 10 |
| 7 | 26132 | 104031 | 307333 | 522703 | 143967 | 492078 | 302967 | 193061 | 74115 | 61229 | 55189 | 10 | 12 | 13 | 13 | 12 | 13 | 13 | 12 | 11 | 11 | 11 |
| 8 | 29766 | 66570 | 172783 | 293719 | 107126 | 365840 | 250341 | 139124 | 52052 | 47573 | 38229 | 10 | 11 | 12 | 13 | 12 | 13 | 12 | 12 | 11 | 11 | 11 |
| 9 | 56075 | 53159 | 155477 | 276672 | 77861 | 271916 | 212318 | 121042 | 44615 | 42478 | 32258 | 11 | 11 | 12 | 13 | 11 | 13 | 12 | 12 | 11 | 11 | 10 |
| 10 | 44875 | 46893 | 130148 | 232122 | 60022 | 173486 | 160137 | 94225 | 34264 | 35150 | 25716 | 11 | 11 | 12 | 12 | 11 | 12 | 12 | 11 | 10 | 10 | 10 |
| 11 | 14019 | 15289 | 42521 | 78588 | 46079 | 69396 | 63025 | 36078 | 12999 | 13297 | 9560 | 10 | 10 | 11 | 11 | 11 | 11 | 11 | 10 | 9 | 9 | 9 |
| 12 | 32359 | 21178 | 61350 | 114600 | 40468 | 40968 | 41490 | 24895 | 9114 | 9132 | 7564 | 10 | 10 | 11 | 12 | 11 | 11 | 11 | 10 | 9 | 9 | 9 |
| 13 | 4848 | 11854 | 39609 | 59932 | 24352 | 58888 | 59380 | 36309 | 13362 | 13774 | 10922 | 8 | 9 | 11 | 11 | 10 | 11 | 11 | 10 | 10 | 10 | 9 |
| 14 | 16837 | 13570 | 31569 | 59060 | 19724 | 30277 | 30355 | 19064 | 7152 | 6682 | 5924 | 10 | 10 | 10 | 11 | 10 | 10 | 10 | 10 | 9 | 9 | 9 |
| 15+ | 109481 | 112947 | 196967 | 349320 | 157707 | 217260 | 239366 | 150688 | 59139 | 49589 | 40797 | 12 | 12 | 12 | 13 | 12 | 12 | 12 | 12 | 11 | 11 | 11 |
| Z (age 7-14) |  |  |  |  |  |  |  |  |  |  |  | 0.17 | 0.33 | 0.36 | 0.33 | 0.29 | 0.45 | 0.36 | 0.37 | 0.31 | 0.31 | 0.33 |
| $F(Z M)$, where $M=0.16$ |  |  |  |  |  |  |  |  |  |  |  | 0.01 | 0.17 | 0.2 | 0.17 | 0.13 | 0.29 | 0.2 | 0.21 | 0.15 | 0.15 | 0.17 |
| Catches (t) |  |  |  |  |  |  |  |  |  |  |  | 21576 | 34751 | 90370 | 144047 | 37096 | 87355 | 75409 | 45231 | 17766 | 19315 | 17388 |
| Correlation coefficient landings vs F . |  |  |  |  |  |  |  |  |  |  |  | 0.46 |  |  |  |  |  |  |  |  |  |  |

Table 3.6.5.1. Boarfish in ICES Subareas 27.6, 7, 8. Estimates of total stock biomass and F. 2018 catch data are not available thus the corresponding $F$ estimate is not available.

| YEAR | TSB.2.5 | TSB.50 | TSB.97.5 | F.2.5 | F.50 | F.97.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 106692 | 207600 | 520907 |  |  |  |
| 1992 | 176295 | 323600 | 794332 |  |  |  |
| 1993 | 227100 | 411950 | 987700 |  |  |  |
| 1994 | 269800 | 506750 | 1231000 |  |  |  |
| 1995 | 218480 | 405650 | 981522 |  |  |  |
| 1996 | 220597 | 412100 | 1005000 |  |  |  |
| 1997 | 198397 | 355800 | 867045 |  |  |  |
| 1998 | 264787 | 483650 | 1172025 |  |  |  |
| 1999 | 197797 | 356500 | 866605 |  |  |  |
| 2000 | 161397 | 293200 | 715922 |  |  |  |
| 2001 | 177900 | 313800 | 765800 |  |  |  |
| 2002 | 156300 | 276850 | 668782 |  |  |  |
| 2003 | 138197 | 241700 | 578712 | 0.00 | 0.00 | 0.00 |
| 2004 | 200797 | 354250 | 843325 | 0.00 | 0.00 | 0.00 |
| 2005 | 193800 | 336700 | 807927 | 0.02 | 0.05 | 0.08 |
| 2006 | 226800 | 395400 | 947322 | 0.01 | 0.01 | 0.03 |
| 2007 | 188597 | 325550 | 775912 | 0.01 | 0.02 | 0.03 |
| 2008 | 235892 | 400800 | 955522 | 0.01 | 0.02 | 0.03 |
| 2009 | 238200 | 404750 | 940012 | 0.03 | 0.07 | 0.11 |
| 2010 | 368597 | 627600 | 1477000 | 0.04 | 0.09 | 0.15 |
| 2011 | 332500 | 566200 | 1355150 | 0.10 | 0.22 | 0.38 |
| 2012 | 494497 | 811700 | 1887000 | 0.10 | 0.23 | 0.39 |
| 2013 | 347592 | 584500 | 1390075 | 0.03 | 0.07 | 0.11 |
| 2014 | 156700 | 261350 | 621905 | 0.05 | 0.11 | 0.18 |
| 2015 | 185597 | 313200 | 742927 | 0.05 | 0.13 | 0.22 |
| 2016 | 124400 | 212700 | 499207 | 0.07 | 0.17 | 0.29 |
| 2017 | 224695 | 384250 | 907027 | 0.02 | 0.06 | 0.10 |
| 2018 | 146800 | 272500 | 655515 | NA | NA | NA |
|  |  |  |  |  |  |  |



Figure 3.1. Boarfish in ICES Subareas 4, 27.6, 7, 8 and 9. Distribution of boarfish in the NE Atlantic area based on presence and absence in IBTS surveys (all years).


Figure 3.1.2.1. Boarfish in ICES Subareas 27.6, 7, 8. Combined Irish boarfish landings 2003-2017 by ICES rectangle (Above). Irish boarfish landings 2017 by ICES rectangle (Below).


Figure 3.2.1.1. Boarfish in ICES Subareas 27.6, 7, 8. Catch numbers-at-age standardised by yearly mean. $15+$ is the plus group.


Figure 3.3.1.1. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish acoustic survey track and haul positions from acoustic survey 2011-2018. Red circles represent 'definitely' boarfish, green: 'probably boarfish', blue: 'boarfish mix' (all included in the biomass estimate).


Figure 3.3.2.1. Boarfish in ICES Subareas 27.6, 7, 8. The haul positions of bottom trawl surveys analysed as an index for boarfish abundance. Note the Portuguese Groundfish survey included here was not included in the 2018 assessment.


Figure 3.3.2.2. Boarfish in ICES Subareas 27.6, 7, 8. Distribution of boarfish in the NE Atlantic showing proposed management area.


Figure 3.3.2.3. Boarfish in ICES Subareas 27.6, 7, 8. CPUE in number per 30 minute haul of boarfish per rectangle in the western IBTS survey 1982 to 2017.


Figure 3.5.1. Boarfish in ICES Subareas 27.6, 7, 8. Recruitment-at-age 1, from various IBTS.


Figure 3.5.2. Boarfish in ICES Subareas 27.6, 7, 8. Recruitment-at-ages 1-5, from various IBTS.


Figure 3.6.1.1. Boarfish in ICES Subareas 27.6, 7, 8. Abundance-at-age in constituent western IBTS Yearly mean standardised abundance-at-age.


Figure 3.6.1.2. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish IBTS survey CPUE fitted delta-lognormal mean (solid line) and 95\% credible intervals (grey region).


Figure 3.6.1.3. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish IBTS survey CPUE data (grey points) and fitted delta-lognormal mean (solid line) and $95 \%$ credible intervals (dashed lines).


Figure 3.6.1.4. Boarfish in ICES Subareas 27.6, 7, 8. Diagnostics from the positive component of the delta-lognormal fits


Figure 3.6.1.5. Boarfish in ICES Subareas 27.6, 7, 8. Pair-wise correlation between the annual mean survey indices.


Figure 3.6.1.6. Boarfish in ICES Subareas 27.6, 7, 8. Weighted correlation between the annual mean survey indices. Correlations are weighted by the sum of the pair-wise variances.


Figure 3.6.3.1. Boarfish in ICES Subareas 27.6, 7, 8. Parameters for final run converged with good mixing of the chains.


Figure 3.6.3.2. Boarfish in ICES Subareas 27.6, 7, 8. Rhat values lower than 1.1 indicating convergence.


Figure 3.6.3.3. Boarfish in ICES Subareas 27.6, 7, 8. MCMC chain autocorrelation for final run.


Figure 3.6.3.4. Boarfish in ICES Subareas 27.6, 7, 8. Residuals around the model fit for the final assessment run.


Figure 3.6.3.5. Boarfish in ICES Subareas 27.6, 7, 8. Prior (red) and posterior (black) distributions of the parameters of the biomass dynamic model.


Figure 3.6.3.6. Boarfish in ICES Subareas 27.6, 7, 8. Trajectories of observed and expected indices for the final assessment run. The stock size over time and a harvest ratio (total catch divided by estimated biomass) are also shown.


Figure 3.6.3.7. Boarfish in ICES Subareas 27.6, 7, 8. Retrospective plot of total stock biomass (above) and fishing mortality (below) from the surplus production model in 2013-2018. Heavy line is current assessment.


Figure 3.6.6.1. Boarfish in ICES Subareas 27.6, 7, 8. Ratios ‘B / MSYBtrigger' and 'F / FMSY' through time and corresponding Kobe plot. Confidence intervals ( 50 and $95 \%$ ) are given for the first two panels, the third displays median estimates only with the pink point representing the first point of the time series and the purple point the last.


Figure 3.9.1.1. Boarfish in ICES Subareas 27.6, 7, 8. Results of exploratory yield per recruit analysis. Beverton and Holt model applied to various fits of the VBGF and for comparison with the VBGF parameters provided by White et al. 2011.


Figure 3.9.1.2. Boarfish in ICES Subareas 27.6, 7, 8. Sensitivity of estimation of F0.1.


Figure 3.12.1. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish samples included in the genetic stock identification study are indicated in green. Population clusters identified by the STRUCTURE analyses are indicated by colour coded circles.

## 4 Herring (Clupea harengus) in subareas 1, 2, and 5, and in divisions 4.a and 14.a, (Northeast Atlantic) (Norwegian Spring Spawning)

### 4.1 ICES advice in 2017

ICES noted that the stock is declining and estimated to be below MSY $B_{\text {trigger }}$ ( 5 million tonnes) in 2017. Since 1998 four large year classes have been produced (1998, 1999, 2002, and 2004). All year classes since 2005 are estimated to be average or small. Fishing mortality has had an overall declining trend since 2010 and was well below Fmsy in 2016.

A long-term management plan agreed by the EU, Faroe Islands, Iceland, Norway and Russia, is operational since 1999. ICES evaluated the plan and concludes that it is in accordance with the precautionary approach. The management plan implies maximum catches of 384197 t in 2018.

### 4.2 The fishery in 2017

### 4.2.1 Description and development of the fisheries

The distribution of the 2017 Norwegian spring-spawning herring (NSSH) fishery for all countries by ICES rectangles per year is shown in Figure 4.2.1.1 and for annual quarter in Figure 4.2.1.2. The 2017 herring fishing pattern was fairly similar to recent years. The fishery began in January on the Norwegian shelf and focused on overwintering, prespawning, spawning and post-spawning fish (Figure 4.2.1.2 quarter 1). In the second quarter, the fishery was insignificant (Figure 4.2.1.2 quarter 2, $0.5 \%$ of total catch). In summer, the fishery had moved into Faroese, Icelandic and Greenlandic waters (Figure 4.2.1.2 quarter 3). In autumn, the fishery had shifted to the overwintering area in the fjords and oceanic areas north of Tromsø and the central part of the Norwegian Sea. In particular, the catches in the international part of the Norwegian Sea were high (Figure 4.2.1.2 quarter 4). The landings in the $1^{\text {st }}$ quarter constituted $22 \%$ of the total landings and the largest proportion of the landings were in the $4^{\text {th }}$ quarter ( $69 \%$ ) which is an increase from 2016, when $52 \%$ of the landings were registered in the $4^{\text {th }}$ quarter.

### 4.3 Stock Description and management units

### 4.3.1 Stock description

A description of the stock is given in the Stock Annex.

### 4.3.2 Changes in migration

Generally, it is not clear what drives the variability in migration of the stock, but the biomass and production of zooplankton are likely factors, as well as feeding competition with other pelagic fish species (e.g. mackerel) and oceanographic conditions (e.g. limitations due to cold areas). Beside environmental factors, the age distribution in the stock will also influence the migration. Changes in migration pattern of NSSH, as well as of other herring stocks, are often linked to large year classes entering the stock initiating a different migration pattern, which subsequent year classes will follow. No large year classes have entered the stock since 2004, although the 2013 year class is estimated to be above average (since 1988) and was in 2018 observed feeding in the north-eastern part of the Norwegian Sea in May and July. In 2017/2018 there was a shift in wintering
areas. While wintering has been observed in fjords west of Tromsø (Norway) for several years, the 2013 year class wintered in fjords farther north (Kvænangen) in 2017/2018 while the older fish seemed to have had an oceanic wintering area. The oldest and largest fish move farthest south and west during feeding, and the older year classes were in May and July 2018 concentrated in the southwestern areas during the feeding season.

### 4.4 Input data

### 4.4.1 Catch data

Catches in tonnes by ICES division, ICES rectangle and quarter in 2017 were available from Denmark, Faroe Islands, Germany, Greenland, Iceland, Ireland, The Netherlands, Norway, Russia, the UK (Scotland), Poland and Sweden. The total working group catch in 2016 was 721566 tonnes (Table 4.4.1.1) compared to the ICES-recommended catch of maximum 437364 tonnes. The majority of the catches ( $91 \%$ ) were taken in area 2.a as in previous years. Samples were not provided by Greenland, the UK or Poland ( $2.5 \%$ of the total catch were taken by these countries). Sampled catches accounted for $95 \%$ of the total catches, which on a similar level assign previous years. The sampling levels of catches in 2017 in total, by country and by ICES division is shown in Table 4.4.1.2, 4.4.1.3 and 4.4.1.4. Catch by nation, ICES division and quarter are shown in Table 4.4.1.5. The software SALLOC (ICES, 1998) was used to calculate total catches in numbers-at-age and mean weight at age representing the total catch. Samples allocated (termed fill-in in SALLOC) to cells (nation, ICES division and quarter) without sampling information are shown in Table 4.4.1.5.

### 4.4.2 Discards

In 2008, the Working Group noted that in this fishery an unaccounted mortality caused by fishing operations and underreporting probably exists (ICES, 2008). It has not been possible to assess the magnitude of these extra removals from the stock, and considering the large catches taken after the recovery of the stock, the relative importance of such additional mortality is probably low. Therefore, no extra mortality to account for these factors has been added since 1994. In previous years, when the stock and the quotas were much smaller, an estimated amount of fish was added to the catches.

The Working Group has not had access to comprehensive data to estimate discards of the herring. Although discarding may occur on this stock, it is considered to be low and a minor problem to the assessment. This is confirmed by estimates from sampling programmes carried out by some EU countries in the Data Collection Framework. Estimates on discarding in 2008 and 2009 of about $2 \%$ in weight were provided for the trawl fishery carried out by the Netherlands. In 2010 and 2012, this métier was sampled by Germany. No discarding of herring was observed ( $0 \%$ ) in either of the two years. An investigation on fisheries induced mortality carried out by IMR with EU partners on fisheries induced and unreported mortality in mackerel and herring fisheries in the North Sea concluded with an estimated level of discarding at around 3\%.

In order to provide information on unaccounted mortality caused by fishing operations in the Norwegian fishery, Ipsos Public Affairs, in cooperation with IMR and the fishing industry, conducted a survey in January/February 2016. The survey was done by phoning skippers and interviewing them. A total of 146 herring skippers participated in the survey, 31 skippers representing the bigger vessel group and 115 skippers representing the smaller vessel group. The data provided an indication that there have been periods
of increased occurrence of net bursting. This was seen especially in the period 20072010. There was, however, no trend in the size of catches where bursting has occurred.

When it comes to slipping, the data showed a steady increase in the percentage that has slipped herring from 2004-2012, and then a significant decline in recent years. The variations in the proportion that have slipped herring were largely driven by the skippers on smaller coastal purse-seiners. Average size of purse-seine hauls slipped seems to be relatively steady over the period. However, the average size of net hauls slipped was lowest in the recent period. An attempt to estimate the level of slipping/bursting (in tonnes) based on these data is planned.

### 4.4.3 Age composition of the catch

The estimated catch-at-age in numbers by years are shown in Table 4.4.3.1. The numbers are calculated using the SALLOC software. In 2017, about $14 \%$ of the catches (in numbers) were taken from both the 2009 year class and the 2013 year class, followed by the 2006 ( $13 \%$ ) and 2011 ( $12 \%$ ) year classes. The 2004 year class still contributes, with $10 \%$ of the catches in 2017.

Catch curves were made on the basis of the international catch-at-age (Figure 4.4.3.1). For comparison, lines corresponding to $\mathrm{Z}=0.3$ are drawn in the background. The big year classes, in the periods of relatively constant effort, show a consistent decline in catch number by cohort, indicating a reasonably good quality of the catch-at-age data. Catch curves for year classes 2005 onwards show a more flat curve than for previous year classes indicating a lower F or a changed exploitation pattern.

### 4.4.4 Weight at age in catch and in the stock

The weight-at-age in the catches in 2017was computed from the sampled catches using SALLOC. Trends in weight-at-age in the catch are presented in Figure 4.4.4.1 and Table 4.4.4.1. The mean weights at age for most of the age groups have generally been increasing in 2010-2013 but levelled off in 2014 and seem to have decreased slightly during the most recent years. A similar pattern is observed in weight-at-age in the stock which is presented in Figure 4.4.4.2 and Table 4.4.4.2. These data have been taken from the survey in the wintering area until 2008. The mean weight at age in the stock for age groups $4-11$ in the years 2009-2017 was derived from samples taken in the fishery in the same area and at the same time as the wintering surveys were conducted in.

### 4.4.5 Maturity-at-age

In 2010 the method for estimating maturity-at-age in the stock assessment of NSSH was changed based on work done by the "workshop on estimation of maturity ogive in Norwegian spring-spawning herring" (WKHERMAT; ICES, 2010a). The method which was adopted by WGWIDE in 2010 (ICES, 2010b) is based on work by Engelhard et al. (2003) and Engelhard and Heino (2004). They developed a method to back-calculate age at maturity for individual herring based on scale measurements, and used this to construct maturity ogives for the year classes 1930-1992.

The NSSH has irregular recruitment pattern with a few large year classes dominating in the stock when it is on a high level. Most of the year classes are, however, relatively small and referred to as "normal" year classes. The back calculation dataset indicates that maturation of the large year classes is slower than for "normal" year classes.

WKHERMAT and WGWIDE considered the dataset derived by back calculation as a suitable potential candidate for use in the assessment because it is conceived in a consistent way over the whole period and can meet standards required in a quality controlled process. However, the back calculation estimates cannot be used for recent years since all year classes have to be fully matured before included. Therefore, assumptions have to be made for recent year classes. For recent year classes, WGWIDE (2010) decided to use average back-calculated maturity for "normal" and "big" year classes, respectively and thereby reducing maturity-at-age for ages 4,5 and 6 when strong year classes enter the spawning stock. The default maturity ogives used for "normal" and "big" year classes are given in the text table below.

| AGE | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | 11 | 12 | 13 | 14 | 15 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| normal <br> ycl | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| strong <br> ycl | 0 | 0 | 0 | 0 | 0.1 | 0.6 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Assumed values should be replaced by back-calculated values in the annual assessments for each year where updated values are available. This was last done in the benchmark assessment in 2016. Therefore, two years (2012 and 2013) could be updated with back-calculated values in the present assessment. Assumed and updated values are shown in figure 4.4.5.1. The maturity ogives used in the present assessment are presented in Table 4.4.5.1.

### 4.4.6 Natural mortality

In this year's assessment, the natural mortality $\mathrm{M}=0.15$ was used for ages 3 and older and $\mathrm{M}=0.9$ was used for ages $0-2$. These levels of M are in accordance to previous years and their justification is provided in the stock annex. Information about deviations from these levels in the time-series, e.g. due to diseases, are also provided in the stock annex.

### 4.4.7 Survey data

The surveys available for the assessment are described in the stock annex. Only two of the available surveys are used in the final assessment and will therefore be dealt with in this section:

1 ) The International Ecosystem Survey in the Nordic Seas (IESNS) in May. The survey covers the entire stock during its migration on the feeding grounds, the adults in the Norwegian Sea and adjacent waters ("Fleet 5") and the juveniles in the Barents Sea ("Fleet 4").

2 ) The Norwegian acoustic survey on the spawning grounds ("Fleet 1") in February.

The cruise reports from the IESNS and spawning survey in 2018 are available as working documents to this report. Both surveys were successfully conducted in 2018.

The abundance estimates from "Fleet 1" are shown in Table 4.4.7.1 and Figure 4.4.7.2; from "Fleet 4" in Table 4.4.7.2 and Figure 4.4.7.1 and "Fleet 5" in Table 4.4.7.3and Figure 4.4.7.1.

Catch curves were made on the basis of the abundance estimates from the surveys "Fleet 1" (Figure 4.4.7.3) and "Fleet 5" (Figure 4.4.7.4). The same arguments are valid for the interpretation of the catch curves from the surveys as from the catches. In 2010,
the number of all age groups decreased suddenly in "Fleet 5" and this is seen as a drop in the catch curves that year. This drop has continued for some of the year classes and the year classes 1998 and 1999 are disappearing faster from the stock than expected. This observed fast reduction in these age classes may also be influenced by the changes in "Fleet 5" catchability, with seemingly higher catchability in years 2006-2009. Like the catch curves from commercial landings, the corresponding curves from "Fleet 5" are also quite flat for year classes 2005 onwards. As "Fleet 1" was not conducted in the years 2009-2014, there is a gap in the catch curves, making it difficult to interpret them.

### 4.4.8 Sampling error in catches and surveys

Sampling errors for Norwegian catch-at-age for the years 2010-2017 is estimated using ECA (Salthaug and Aanes 2015, Hirst et al. 2012). Using the Taylor function (Aanes 2016a) to model the sampling variance of the catches yields a very good fit ( $R_{a d j}^{2}=$ 0.94 ) and using this function to impute missing sampling variances for catch-at-age yields relative standard errors shown in Table 4.4.8.1. It is assumed that the relative standard errors in the total catches are equal to the Norwegian catches (which comprise $\sim 60 \%$ of the total catches). Sampling errors for survey indices are estimated using StoX (http://www.imr.no/forskning/prosjekter/stox/nb-no). For Fleet 1 estimates are available for the years 1988-1989, 1994-1996, 1998-2000, 2005-2008, and 2015-2018, for Fleet 4 estimates of sampling errors are available for 2009-2018, and for Fleet 5 for 20082018. Missing values for sampling variances are imputed using the Taylor function which provides goods fits ( $R_{a d j}^{2}$ 's are $0.94,0.98,0.96$, respectively). The resultant relative standard errors are given in Tables 4.4.8.2-4.4.8.4. Due to the very good fits of the Taylor functions, estimates of relative standard where empirical estimates are available, are also replaced by the model predicted values to reduce potential effects of imprecise estimates of errors.

### 4.4.9 Information from the fishing industry

No information is made available for the working group.

### 4.5 Stock assessment

The first benchmark of the NSSH took place in 2008. The assessment tool TASACS was then chosen to be the standard assessment tool for the stock. The second benchmark took place in 2016 (ICES 2016c) where three assessment models were explored, TASACS, XSAM and one separable model. WKPELA accepted XSAM as the standard assessment tool for the NSSH.

### 4.5.1 XSAM final assessment 2018

The XSAM model is documented in Aanes 2016a and 2016b. XSAM includes the option to utilize the prediction of total catch in the assessment year (typically sum of national quotas) along with the precision of the prediction. This was changed in 2017 as it was found that the model estimated a highly variable and significantly lower compared to the working group's prediction (sum of national quotas). In addition, this caused an abrupt change in the selection pattern from 2017 and onwards. The abrupt change in the selection pattern was not fully understood by the working group, but the effect was less pronounced if not using the catch prediction from the model for 2017. Therefore, it was decided to not utilize the prediction of total catches in 2017 when fitting the model to data (i.e. the assessment) and consequently in the short-term forecast. The same approach is taken in the 2018 assessment, i.e. the catch prediction for 2018 is not included when fitting the model to data. The resulting estimated selection pattern is
gradual (Figure 4.5.1.1) and in line with the current knowledge about the fishery. It is important to notice that this has marginal effect on the assessment, but larger effects on the prediction and short-term forecast.

This year's XSAM assessment was performed with the same model options as in 2017. In summary this means that the model was fitted with time varying selectivity and effort according to $\operatorname{AR}(1)$ models in the model for fishing mortality; the recruitment was modelled as a process with constant mean and variance; the standard errors for all input data were predetermined using sample data (Tables 4.4.8.1-4.4.8.4), but estimating a scaling constant common for all input data to allow additional variability in the input data that is not controlled by sampling. Other details in settings are given in the Stock Annex.

The same input data over the same age ranges was used as in 2017. At the 2016 benchmark, data from 1988 and onwards was used, the considered age-span was 3-12+ with input data catch-at-age, Fleet 1 and Fleet 5 and in WGWIDE 2016 it was decided to start the model at age 2 to enable short-term predictions with reasonable levels of variability. To achieve this, age 2 from Fleet 4, and age 2 in catch-at-age is included in input data. Evaluation of diagnostics including lower ages than 2 and/or other fleets resulted in excluding lower ages than 2 and other fleets for the final assessment. Input data are listed in Table C.1.1 in the Stock Annex.

The parameter estimates are shown in Table 4.5.1.1. For a precise definition of the parameters it is referred to Aanes 2016a in ICES (2016). Note that the variance components $\sigma_{1}^{2}$ (variability in the separable model for F ) and $\sigma_{R}^{2}$ (variability in recruitment) is rather imprecise. The estimate of the scaling constant $h$ is larger than 1 showing that the model adds additional variability on the observation errors than explained by the sampling errors alone.

The catchabilities for all the fleets are on average positively correlated indicating some uncertainty due to a common scaling of all surveys to the total abundances although the correlations in general are small (Figure 4.5.1.2). There is a slight negative correlation between $\sigma_{1}^{2}$ and $\sigma_{2}^{2}$ (variability in the AR process for time varying selectivity) indicating little contrast in data for separating variability in the separable model from variability due to changes in selection pattern. The slopes in the multivariate AR model for time-varying selectivity gradually changes from negative to positive, but is expected as it is imposed due to the sum to zero constraint for the selection (see Aanes 2016a for details).

The weights each datum is given in the model fit (inverse of the sampling variance) is proportional to the empirical weights derived from sampling variances (Tables 4.4.8.14.4.8.4) which shows that the strong year classes in general is given larger weight to the model than weak year classes, and the ordering of the average weights (from high to low) is Catch-at-age, Fleet 5, Fleet 1 and Fleet 4 (Figure 4.5.1.3).

Two types of residuals are considered for this model. The first type is the model prediction (based on all data) vs. the data. In such time-series models, the residuals based on the prediction which uses all data points will be serially correlated although useful as they explain the unexplained part of the model (cf Harvey 1990 p 258).This means that patterns in residuals over time is to be expected and questions the use of e.g. qqplots as an additional diagnostic tools to assess distributional assumptions. To obtain residuals which follow the assumptions about the data in the observation models (e.g. serially uncorrelated) single joint sample residuals are extracted (ICES 2017). In short these are obtained by sampling predicted values from the conditional distribution of
values given the observations. This sample corresponds to a sample from the joint distribution of latent variables and observations. The third approach could have been to extract the one step ahead observation residuals which are standard for diagnostics for regular state-space models (cf Harvey 1990). This is not done here.
The negative residuals tracing the 1983 year class for catch-at-age represents low fishing mortalities examining the type 1 residuals (Figure 4.5.1.4). This effect is less pronounced considering the type 2 residuals. The type 2 residuals are qualitatively comparable with the type 1 residuals but generally display more mixed residuals as predicted by the theory. Otherwise the residuals for catch-at-age appears fairly mixed apart for some serial correlation for age 2 and 3 (which are very low), and some negative residuals for the plus group the most recent years. The residuals for Fleet 1 in 1994 and 2015 for young and old ages are all of the same signs and may appear as year effects. Also note that the residuals for Fleet 1 for ages $10+$ in 2015 and 2016 are all positive (Figure 4.5.1.4) which shows that the abundance indices from Fleet 1 displays a larger stock size over these ages and years compared to the assessment using all input data. However, these data points are given low weights (Figure 4.5.1.3) as they are found imprecise (Tables 4.4.8.1-4.4.8.4). Some serial correlation for residuals for ages 3 and 4 in Fleet 1 can also be detected, but is down weighted by the same reasons. Serial correlation in residuals for age 2 in Fleet 4 can also be detected indicating trends over time in mismatch between estimates and observations of abundance at age 2. Residuals for Fleet 5 appears adequate compared to previous years although some serial correlations can be detected also here.

The residuals for small values are bigger than residuals for the larger values since smaller values in general have higher variances than larger values (Tables 4.4.8.14.4.8.4) (Figure 4.5.1.5).The qq-plots for the standardized residuals show that the distributional assumptions on the observation errors are adequate, except for the smallest and largest values of catch-at-age and indices from Fleet 1. As qq-plots for residuals of type 1 may be questioned (see above) it is noted that qq-plots for residuals of type 2 is more relevant and generally shows a significantly better fit based on a visual inspection compared to using type 1 .

The marginal likelihood and the components for each data source (see Aanes 2016b for details) are profiled over a range of the common scaling factor $h$ for all input data (Figure 4.5.1.6). It is apparent that the optimum of the marginal likelihood is clearly defined. The catch component is decreasing with decreasing values of $h$ indicating that the model puts more weight on the catch component than indicated by the comparing sampling errors for all input data. This is in line with the findings in Aanes (2016a and 2016b) who showed that these types of models tends to put too much weight on the catch data if the weighting is not constrained. However, the likelihood component for the catch is overruled by the information in Fleet 1, 4 and 5 such that the optimum for the marginal likelihood is clearly defined. The point estimates of SSB and F is insensitive to different values of $h$.

The retrospective runs for this model shows estimates which is within the estimated levels of precision (Figure 4.5.1.7). The indices from Fleet 1 indicate, on average, a relatively larger abundance than the indices from Fleet 5 for 2015-2017 which is supported by the positive residuals for ages $9-10+$ (Figure 4.5.1.4). Consequently, the increased estimates of SSB and decreased estimates of F after 2014 is a response to the indices from Fleet 1 which not was conducted in the years 2009-2014. Note that the retrospective estimates are remarkably stable from 2015 and onwards. To illustrate the conflict in data and increased uncertainty in estimates the most recent years, the abundance
indices are scaled to the absolute abundance by the estimated catchabilities. Then the spawning-stock biomass based on each survey index is calculated using the stock weights at age and proportion mature at age (Figure 4.5.1.8). Here we see a fairly good temporal match between the model estimate of SSB and the survey SSBs except for the years 2015 and 2016 for Fleet 1, which display a significantly faster reduction in the stock compared to Fleet 5 which shows a more flat trend in the same years. It is worth noticing that although the point estimate of SSB based on Fleet 1 appear very much higher than Fleet 5 in 2015 and 2016, the uncertainty in the estimates are very high, such that the respective estimates do not appear as significantly different. However, the effect on the final assessment is to lift the point estimate of SSB and increase the uncertainty which is in accordance with the data used (Figure 4.5.1.9).

The final assessment results are shown in Figure 4.5.1.9. The estimates of fishing mortality for 2017 is rather high, as a response to the high catch in 2017 with a point estimate of 0.174 although the estimate is rather imprecise since the $95 \%$ confidence interval ranges from $0.123-0.224$. The spawning stock shows a declining trend since 2009, and the $95 \%$ confidence interval of the stock level in 2018 ranges from $\sim 3.1$ to $\sim 4.6$ million tonnes which barely envelopes $\mathrm{B}_{\mathrm{mp}}=3.184$ million tonnes, such that the probability of the stock being above $\mathrm{B}_{\mathrm{lim}}=2.5$ million tonnes is high. Note the rather large uncertainty in the absolute levels since the peak in 2009 with the further increase in the most recent years. This high uncertainty is a result of the conflicting signals in data concerning the degree of decrease in the stock over this time period.

The final results of the assessment are also presented in Tables 4.5.1.2 (stock in numbers), 4.5.1.3 (fishing mortality) and Table 4.5.1.4 is the summary table of the assessment.

### 4.5.2 Exploratory assessments

### 4.5.2.1 TASACS

TASACS was run according to the benchmark in 2008 using the VPA population model in the TASACS toolbox with the same model options as the benchmark (see Stock Annex). The information used in the TASACS run is catch data and survey data from eight surveys. The analysis was restricted to the years 1988-2018. The model was run with catch data from 1988 to 2017, and projected forwards through 2018 assuming Fs in 2018 equal to those in 2017, to include survey data from 2018. The larval survey (SSB fleet) was discontinued in 2017 and no new information is therefore available from this survey.

The model fit to the tuning data is shown with Q-Q plots in Figure 4.5.2.1.1. Surveys 1, 2,3 and 7 seem to fit rather well to the assumed linear relationship in the TASACS model, but surveys 4,6 and 8 have rather poor fit. Since 2016 the TASACS run Q-Q plots for fleet 5 shows a poorer fit compared to earlier assessments. This is mainly caused by a change in estimated catchability.

Particularly Survey 8 (larval survey) seems to have a poor fit. This can also be seen as a block of positive residuals for this survey in later years (Figure 4.5.2.1.2). The residual plot for survey 5 (IESNS) also shows some pattern with consecutive series of negative and positive residuals indicating year-effects.

The results from TASACS are compared to those from XSAM and TISVPA in Figure 4.5.2.1.3.The time-series of SSB show similar trends for XSAM and TASACS while TISVPA do not show the same downward trend in the later period. For most of the years,
the estimates from TASACS and TISVPA are mostly within the confidence limits estimated by XSAM. The SSB on 1 January 2018 is estimated by TASACS to be 3.693 million tonnes, which is lower than the estimated value from TISVPA but close to the point estimate from XSAM.

### 4.5.2.2 TISVPA

The TISVPA model was applied using the catch-at-age data with range from 0 to $15+$ and data from three surveys (Survey 1, 4 and 5). No data points were down-weighted. Two-parametric selection pattern used in the model revealed some obvious peculiarities in the interaction between the stock and the fishery.
Rather clear signals about the stock biomass in 2018 were obtained from just catch-atage and surveys 1,4 and 5 . Catch-at-age and Survey 1 data, as well as the overall objective function of the model, indicate the SSB value in 2018 about 4.7 million tonnes (see WD 12). Surveys 4 and 5 indicate the SSB value about 6 and 4 million tonnes respectively.
The results from TISVPA are compared to those from XSAM and TASACS in Figure 4.5.2.1.3.

### 4.6 NSSH reference points

ICES last reviewed the reference points of Norwegian spring spawning herring in April 2018. ICES concluded that Blim should remain unchanged at 2.5 million tonnes and MSYB trigger $=B_{\text {pa }}$ was estimated at 3.184 million tonnes. FMSY was estimated at 0.102 , but during an ongoing work on Management Strategy Evaluation Fmsy has been revisited, because issues were found with numerical instability and settings when $\mathrm{F}_{\text {MSY }}=0.102$ was set. Therefore Fmsy is currently being re-estimated.

### 4.6.1 PA reference points

The PA reference points for the stock were last estimated by WKNSSHREF in 2018. The
 3.184 million tonnes and $\mathrm{F}_{\mathrm{pa}}=0.182$. $\mathrm{F}_{\mathrm{pa}}$ is presently being revisited in WKNSSHMSE.

### 4.6.2 MSY reference points

The MSY reference points were evaluated by WKNSSHREF in 2018. In the ICES MSY framework Bpa is proposed/adopted as the default trigger biomass Btrigger and was estimated at 3.184 million tonnes. Fmsy is currently being revisited by WKNSSHMSE.

### 4.6.3 Management reference points

In the current management plan the Coastal States have agreed a target reference point defined at $F_{\text {target }}=0.125$ when the stock is above $B_{p a}$. If the $S S B$ is below $B_{p a}$, a linear reduction in the fishing mortality rate will be applied from 0.125 at $\mathrm{B}_{\mathrm{pa}}$ to 0.05 at $\mathrm{Blim}_{\text {. }}$

There is ongoing work (WKNSSHMSE) to answer a request from the Coastal States on updated Management Strategy.

### 4.7 State of the stock

The SSB on 1 January 2018 is estimated by XSAM to be 3.826 million tonnes which is above $B_{p a}(3.184$ million $t)$. The stock is declining and the SSB time-series from the 2018 assessment is in line with the SSB time-series from the 2017 assessment. In the last 15 years, five large year classes have been produced (1998, 1999, 2002, 2003, and 2004). The 2005 to 2015 year classes are estimated to be average or small, however, the 2016 year class is estimated to be well above average (from 1988). Fishing mortality in 2017 is estimated to be 0.174 which is above the management plan $F$ that was used to give advice for 2017. A new management plan is being developed for the 2019 advisory year.

### 4.8 NSSH Catch predictions for 2018

### 4.8.1 Input data for the forecast

Forecasting was conducted using XSAM according to the method described in the Stock Annex and by Aanes (2016c). WGWIDE 2016 decided to use the point estimates from this forecast as basis for the advice. In short the forecast is made by applying the point estimates of the stock status as input to set TAC, then based on the TAC a stochastic forecast were performed to determine levels of precision in the forecast. Table 4.8.1.1 list the point estimates of the starting values for the forecast. The input stock numbers-at-age 2 and older were taken from the final assessment. The catch weight-at-age, used in the forecast, is the average of the observed catch weights over the last 3 years (2015-2017).
For the weight-at-age in the stock, the values for 2018 were obtained from the commercial fisheries in the wintering areas in January. For the years 2019 and 2020 the average of the last 3 years (2016-2018) was used.

Standard values for natural mortality were used. Maturity-at-age was based on the information presented in Section 4.4.5.

The exploitation pattern used in the forecast is taken from the predictions made by the model (see Aanes 2016c for details). The resultant mean annual exploitation pattern is shown in Figure 4.8.1.1 and displays a shift towards older fish in the recent years and further in the prediction. Prediction of recruitment at age 2 is obtained by the model with a mean that in practice represents the long term (1988-2018) estimated mean recruitment (back-transformed mean at log scale) and variance the corresponding recruitment variability over the period. Forecasted values of recruits are highly imprecise but have little influence on the short-term forecast of SSB as the herring starts to mature at age 4 .

The average fishing mortality defined as the average over the ages 5 to 12 is weighted over the population numbers in the relevant year

$$
\bar{F}_{y}=\sum_{a=5}^{12} N_{a, y} F_{a, y} / \sum_{a=5}^{12} N_{a, y}
$$

where $F_{a, y}$ and $N_{a, y}$ are fishing mortalities and numbers by age and year. This procedure is in accordance with previous years for this stock but the age range is shifted from 5-11 to 5-12.

There was no agreement of a TAC for 2018. To obtain an estimate of the total catch to be used as input for the catch-constraint projections for 2019 , the sum of the unilateral
quotas was used. In total, the expected outtake from the stock in 2018 amounts to 546 448 tonnes. F in 2018 is estimated by XSAM based on this catch.

### 4.8.2 Results of the forecast

The Management Options Table with the results of the forecast is presented in Table 4.8.2.1. Assuming a total catch of 546448 tonnes is taken in 2018, it is expected that the SSB will increase from 3.826 million tonnes ( $95 \%$ confidence interval 3.065 to 4.587 million tonnes) on 1 January in 2018 to 3.859 million tonnes in 2019 ( $95 \%$ confidence interval 3.069 to 4.866 million tonnes). The $95 \%$ confidence interval for weighted $F$ over ages $5-11$ in 2018 ranges from 0.03 to 0.275 with a mean of 0.117 , while the corresponding values for ages 5-12 are $0.035,0.280$ and 0.125 , respectively.

### 4.9 Comparison with previous assessment

A comparison between the assessments 2008-2017 is shown in Figure 4.9.1. In the years 2008 - 2015 the assessments were made with TASACS, whereas since 2016 XSAM has been applied, as accepted by WKPELA 2016. With the change of the assessment tool in 2016 the age of the recruitment changed from 0 to 2 and the age span in the reference F changed from $5-14$ to $5-11$. In WKNSSHREF (2018) this was further changed to 5-12.

The table below shows the SSB (thousand tonnes) on 1 January in 2017 and weighted F in 2016 as estimated in 2017 and 2018.

|  | ICES 2017 |  | WG 2018 | \%DIFFERENCE |
| :--- | ---: | :--- | :--- | :--- |
| SSB(2017) | 4131 | 4235 | $2.5 \%$ |  |
| Weighted F $(2016)^{*}$ | 0.084 | 0.092 |  |  |

*F in the 2017 assessment was based on the age span 5-11 and therefore not directly comparable to the F in the 2018 assessment which was based on the age span 5-12.

### 4.10 Management plans and evaluations

The long-term management plan of Norwegian spring spawning herring aims for exploitation at a target fishing mortality below $\mathrm{F}_{\mathrm{pa}}$ and is considered by ICES in accordance with the precautionary approach (WKBWNSSH, ICES, 2013d). The management plan in use contains the following elements:

Every effort shall be made to maintain a level of Spawning-stock biomass (SSB) greater than the critical level ( $\mathrm{Blim}_{\mathrm{lim}}$ ) of 2500000 t .

For 2012 and subsequent years, the Parties agreed to restrict their fishing on the basis of a TAC consistent with a fishing mortality rate of less than 0.125 for appropriate age groups as defined by ICES, unless future scientific advice requires modification of this fishing mortality rate.

Should the SSB fall below a reference point of $5000000 t\left(\mathrm{~B}_{\mathrm{pa}}\right)$, the fishing mortality rate, referred under Paragraph 2, shall be adapted in the light of scientific estimates of the conditions then prevailing to ensure a safe and rapid recovery of the SSB to a level in excess of 5000000 t . The basis for such adaptation should be at least a linear reduction in the fishing mortality rate from 0.125 at $B_{p a}(5000000 t)$ to 0.05 at $B_{\lim }(2500000 t)$.

The Parties shall, as appropriate, review and revise these management measures and strategies on the basis of any new advice provided by ICES.

A brief history of it is in the stock annex. In general, the stock has been managed in compliance with the management plan.

There is ongoing work to answer a request from the Coastal States on updated Management Strategy, which will be based on the new MSY reference points.

### 4.11 Management considerations

Perception of the stock has not changed since last year's assessment (estimated SSB in 2017 is $2.5 \%$ higher in this year's assessment). Results of exploratory runs by other models match with those of XSAM.
Historically, the size of the stock has shown large variations and dependency on the irregular occurrence of very strong year classes. Between 1998 and 2004 the stock produced several strong year classes which lead to an increase in SSB until 2009. Since then, SSB has declined due to absence of strong year classes after 2004, but the 2016 year class is estimated to be above average (since 1988).
Since 1999 catches have been regulated through an agreed management plan, which is considered to be precautionary. However, since 2013, a lack of agreement by the Coastal States on their share in the TAC has led to unilaterally set quotas which together are higher than the TAC indicated by the management plan resulting in steeper reduction in the SSB than otherwise.

At present work on management strategy evaluation is ongoing and a new management strategy is expected to be in place for the advisory year 2019.

### 4.12 Ecosystem considerations

NSS herring juveniles and adults are an important part of the ecosystems in the Barents Sea, along the Norwegian coast, in the Norwegian Sea and in adjoining waters. This refers both to predation on zooplankton by herring and herring being a food resource to higher trophic levels (e.g. cod, saithe, seabirds, and marine mammals). The predation intensity of and on herring have seasonal, spatial and temporal variation as a consequence of variation in migration pattern, prey density, stock size, size of year classes and stock sizes of competing stocks for resources and predators. Recent features of some of these ecosystem factors of relevance for the stock are summarized below.

- The stock's more westerly feeding distribution in recent years (ICES 2017a; 2017b) might be due to better feeding opportunities there or a response to feeding competition with mackerel but the consequence is a less spatial overlap of herring and mackerel in Norwegian Sea and adjoining waters since around 2014 (Nøttestad et al., 2014; ICES, 2015b; 2016b; 2017b).
- Where herring and mackerel overlap spatially they compete for food to some extent (Bachiller et al., 2015; Debes et al., 2012; Langøy et al., 2012; Óskarsson et al., 2015) but studies showing mackerel being more effective feeder might indicate that the herring is forced to the western and northern fringe of Norwegian Sea, although higher zooplankton biomass there could also attract the herring (Nøttestad et al., 2014; ICES, 2015b; 2016b).
- Results of stomach analyses of mackerel on the Norwegian coastal shelf (between about $66^{\circ} \mathrm{N}$ and $69^{\circ} \mathrm{N}$ ) suggest that mackerel fed opportunistically on herring larvae, and that predation pressure therefore largely depends on the degree of overlap in time and space (Skaret et al., 2015).
- The 2013 year class of herring is the strongest since the 2004 year class. In the May survey it was found both in the north eastern and in the central part of the Norwegian Sea.
- Herring growth (i.e. length-at-age) varied over the period 1994-2015 and was negatively related to stock size (Homrum et al., 2016), which indicates interaction between fish density and prey availability.
- Following a maximum in zooplankton biomass during the early 2000s the biomass declined with a minimum in 2006. From 2010, the trend turned to an increase and reached the long-term mean in 2014. Zooplankton biomass dropped again in 2015, but has been increasing since then. Interestingly, all the areas, excluding east of Iceland and on few occasions Jan Mayen AF (Figure 6.2), show parallel changes in zooplankton biomass.
- The Subpolar gyre, which has been in a weak state since mid 1990's, has been strengthening during the last three years. If this trend continues, we should expect increased levels of silicate entering the Norwegian Sea over the coming years and consequently a reversal in the declining trend of silicate observed in the Norwegian Sea since 1990. Increasing silicate concentrations are expected to affect growth of silicate demanding phytoplankton, which again will affect zooplankton grazing (ICES, 2018a, and references therein).
- The temperatures of the inflowing Atlantic water were in 2017 above the long-term means (1981-2010) for the whole region. The salinity in the Atlantic Water was below the long-term means in the south and close to or higher than the normal in the north. The heat content increased in the North and Norwegian Seas and it was record-high in the Norwegian Sea. In the Barents Sea the ice cover during 2017 was below the long-term mean during the whole year (ICES, 2018b).


### 4.13 Changes in fishing patterns

The fishery for Norwegian spring spawning herring has generally been described as progressing clockwise in the Nordic Seas as the year progresses. In the recent years (after $\sim 2013$ ) this pattern has changed, because there has been an extended fishery in the south and southwestern areas in the Norwegian Sea in the $3^{\text {rd }}$ and $4^{\text {th }}$ quarters $(8 \%$ and $69 \%$ respectively in 2017), and thus almost $3 / 4$ 's of the herring catch was taken in the last quarter of 2017. The majority of the catches in the $4^{\text {th }}$ quarter are now taken in the central parts of the Norwegian Sea, whereas in the preceding years there was a more significant fishery in northeastern areas (outside northern Norway and southwest of the Bear Island). This change in migration resulted in late arrival at the Norwegian coast for this part of the stock during the winter 2017/2018. The Norwegian coastal fleet (smaller vessel that cannot go that far offshore) could therefore not access this herring during the winter fishery and targeted younger fish (mostly of the 2013 and 2014 year classes) which overwintered in Norwegian fjords.

### 4.14 Recommendation

In the IESNS survey other herring stocks (e.g. Icelandic summer spawning herring and North Sea herring) are found in the boundary regions of the survey area. WGWIDE recommends that WGIPS initiates work to distinguish between herring stocks on the individual level as well as to provide abundance indices by stock.

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### 4.16 Tables

Table 4.4.1.1 Total landings (ICES estimate) of Norwegian spring-spawning herring (tons) since 1972. Data provided by Working Group members.

| Year | Norway | $\begin{aligned} & \text { USSR/ } \\ & \text { RussIA } \end{aligned}$ | Denmark | Faroes | ICELAND | Ireland | Netherlands | Greenland | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 13161 | - | - | - | - | - | - | - | - | - | - | - | - | 13161 |
| 1973 | 7017 | - | - | - | - | - | - | - | - | - | - | - | - | 7017 |
| 1974 | 7619 | - | - | - | - | - | - | - | - | - | - | - | - | 7619 |
| 1975 | 13713 | - | - | - | - | - | - | - | - | - | - | - | - | 13713 |
| 1976 | 10436 | - | - | - | - | - | - | - | - | - | - | - | - | 10436 |
| 1977 | 22706 | - | - | - | - | - | - | - | - | - | - | - | - | 22706 |
| 1978 | 19824 | - | - | - | - | - | - | - | - | - | - | - | - | 19824 |
| 1979 | 12864 | - | - | - | - | - | - | - | - | - | - | - | - | 12864 |
| 1980 | 18577 | - | - | - | - | - | - | - | - | - | - | - | - | 18577 |
| 1981 | 13736 | - | - | - | - | - | - | - | - | - | - | - | - | 13736 |
| 1982 | 16655 | - | - | - | - | - | - | - | - | - | - | - | - | 16655 |
| 1983 | 23054 | - | - | - | - | - | - | - | - | - | - | - | - | 23054 |
| 1984 | 53532 | - | - | - | - | - | - | - | - | - | - | - | - | 53532 |
| 1985 | 167272 | 2600 | - | - | - | - | - | - | - | - | - | - | - | 169872 |
| 1986 | 199256 | 26000 | - | - | - | - | - | - | - | - | - | - | - | 225256 |
| 1987 | 108417 | 18889 | - | - | - | - | - | - | - | - | - | - | - | 127306 |
| 1988 | 115076 | 20225 | - | - | - | - | - | - | - | - | - | - | - | 135301 |
| 1989 | 88707 | 15123 | - | - | - | - | - | - | - | - | - | - | - | 103830 |
| 1990 | 74604 | 11807 | - | - | - | - | - | - | - | - | - | - | - | 86411 |
| 1991 | 73683 | 11000 | - | - | - | - | - | - | - | - | - | - | - | 84683 |
| 1992 | 91111 | 13337 | - | - | - | - | - | - | - | - | - | - | - | 104448 |
| 1993 | 199771 | 32645 | - | - | - | - | - | - | - | - | - | - | - | 232457 |
| 1994 | 380771 | 74400 | - | 2911 | 21146 | - | - | - | - | - | - | - | - | 479228 |
| 1995 | 529838 | 101987 | 30577 | 57084 | 174109 | - | 7969 | 2500 | 881 | 556 | - | - | - | 905501 |
| 1996 | 699161 | 119290 | 60681 | 52788 | 164957 | 19541 | 19664 | - | 46131 | 11978 | - | - | 22424 | 1220283 |
| 1997 | 860963 | 168900 | 44292 | 59987 | 220154 | 11179 | 8694 | - | 25149 | 6190 | 1500 | - | 19499 | 1426507 |


| Year | Norway | USSR/ <br> Russia | Denmark | Faroes | ICELAND | Ireland | Netherlands | Greenland | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 743925 | 124049 | 35519 | 68136 | 197789 | 2437 | 12827 | - | 15971 | 7003 | 605 | - | 14863 | 1223131 |
| 1999 | 740640 | 157328 | 37010 | 55527 | 203381 | 2412 | 5871 | - | 19207 | - | - | - | 14057 | 1235433 |
| 2000 | 713500 | 163261 | 34968 | 68625 | 186035 | 8939 | - | - | 14096 | 3298 | - | - | 14749 | 1207201 |
| 2001 | 495036 | 109054 | 24038 | 34170 | 77693 | 6070 | 6439 | - | 12230 | 1588 | - | - | 9818 | 766136 |
| 2002 | 487233 | 113763 | 18998 | 32302 | 127197 | 1699 | 9392 | - | 3482 | 3017 | - | 1226 | 9486 | 807795 |
| 2003* | 477573 | 122846 | 14144 | 27943 | 117910 | 1400 | 8678 | - | 9214 | 3371 | - | - | 6431 | 789510 |
| 2004 | 477076 | 115876 | 23111 | 42771 | 102787 | 11 | 17369 | - | 1869 | 4810 | 400 | - | 7986 | 794066 |
| 2005 | 580804 | 132099 | 28368 | 65071 | 156467 | - | 21517 | - | - | 17676 | 0 | 561 | 680 | 1003243 |
| 2006* | 567237 | 120836 | 18449 | 63137 | 157474 | 4693 | 11625 | - | 12523 | 9958 | 80 | - | 2946 | 968958 |
| 2007 | 779089 | 162434 | 22911 | 64251 | 173621 | 6411 | 29764 | 4897 | 13244 | 6038 | 0 | 4333 | 0 | 1266993 |
| 2008 | 961603 | 193119 | 31128 | 74261 | 217602 | 7903 | 28155 | 3810 | 19737 | 8338 | 0 | 0 | 0 | 1545656 |
| 2009 | 1016675 | 210105 | 32320 | 85098 | 265479 | 10014 | 24021 | 3730 | 25477 | 14452 | 0 | 0 | 0 | 1687371 |
| 2010 | 871113 | 199472 | 26792 | 80281 | 205864 | 8061 | 26695 | 3453 | 24151 | 11133 | 0 | 0 | 0 | 1457015 |
| 2011 | 572641 | 144428 | 26740 | 53271 | 151074 | 5727 | 8348 | 3426 | 14045 | 13296 | 0 | 0 | 0 | 992997 |
| 2012 | 491005 | 118595 | 21754 | 36190 | 120956 | 4813 | 6237 | 1490 | 12310 | 11945 | 0 | 0 | 705 | 826000 |
| 2013 | 359458 | 78521 | 17160 | 105038 | 90729 | 3815 | 5626 | 11788 | 8342 | 4244 | 0 | 0 | 23 | 684743 |
| 2014 | 263253 | 60292 | 12513 | 38529 | 58828 | 706 | 9175 | 13108 | 4233 | 669 | 0 | 0 | 0 | 461306 |
| 2015 | 176321 | 45853 | 9105 | 33031 | 42625 | 1400 | 5255 | 12434 | 55 | 2660 | 0 | 0 | 0 | 328740 |
| 2016 | 197501 | 50455 | 10384 | 44727 | 50418 | 2048 | 3519 | 17508 | 4031 | 2582 | 0 | 0 | 0 | 383174 |
| 2017 | 389383 | 91118 | 19037 | 98170 | 90400 | 3495 | 6679 | 12569 | 4358 | 5201 | 0 | 1 | 1155 | 721566 |

*In 2003 the Norwegian catches were raised of 39433 to account for changes in percentages of water content

Table 4.4.1.2 Norwegian spring-spawning herring. Sampling coverage by year.

| Year | TOTAL CATCH | \% catch covered by SAMPLING PROGRAMME | No. SAMPLES | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 1207201 | 86 | 389 | 55956 | 10901 |
| 2001 | 766136 | 86 | 442 | 70005 | 11234 |
| 2002 | 807795 | 88 | 184 | 39332 | 5405 |
| 2003 | 789510 | 71 | 380 | 34711 | 11352 |
| 2004 | 794066 | 79 | 503 | 48784 | 13169 |
| 2005 | 1003243 | 86 | 459 | 49273 | 14112 |
| 2006 | 968958 | 93 | 631 | 94574 | 9862 |
| 2007 | 1266993 | 94 | 476 | 56383 | 14661 |
| 2008 | 1545656 | 94 | 722 | 81609 | 31438 |
| 2009 | 1686928 | 94 | 663 | 65536 | 12265 |
| 2010 | 1457015 | 91 | 1258 | 124071 | 12377 |
| 2011 | 992.997 | 95 | 766 | 79360 | 10744 |
| 2012 | 825.999 | 93 | 649 | 59327 | 14768 |
| 2013 | 684.743 | 91 | 402 | 33169 | 11431 |
| 2014 | 461.306 | 89 | 229 | 18370 | 5813 |
| 2015 | 328.739 | 92 | 177 | 25156 | 5039 |
| 2016 | 383.174 | 91 | 203 | 39120 | 5892 |
| 2017 | 721566 | 95 | 335 | 31755 | 7241 |

Table 4.4.1.3 Norwegian spring-spawning herring. Sampling coverage by country in 2017.

| COUNTRY | OFFICIAL CATCH | \% CATCH covered by SAMPLING programme | NO. <br> SAMPLES | NO. MEASURED | NO. AGED |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 19037.4 | 74 | 5 | 704 | 140 |
| Faroe Islands | 98170.3 | 94 | 13 | 806 | 666 |
| Germany | 5201.1 | 99 | 5 | 321 | 321 |
| Greenland | 12569 | 0 | 0 | 0 | 0 |
| Iceland | 90400 | 100 | 90 | 2164 | 2008 |
| Ireland | 3494.7 | 100 | 2 | 91 | 76 |
| Norway | 389383.5 | 99 | 94 | 2222 | 2222 |
| Poland | 0.7 | 0 | 0 | 0 | 0 |
| The Netherlands | 6678.8 | 94 | 29 | 1854 | 725 |
| UK_Scotland | 4358 | 0 | 0 | 0 | 0 |
| Sweden | 1155 | 0 | 0 | 0 | 0 |
| Russia | 91118 | 99 | 97 | 23595 | 1083 |
| Total for Stock | 721566 | 95 | 335 | 31755 | 7241 |

Table 4.4.1.4 Norwegian spring-spawning herring. Sampling coverage by ICES Division in 2017.

| Area | Official Catch | No <br> SAMPLES | $\begin{gathered} \text { No } \\ \text { AGED } \end{gathered}$ | No Measured | No Aged/ 1000 TONNES | No Measured/ 1000 TONNES |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2.a | 660042.9 | 278 | 5990 | 30414 | 9 | 46 |
| 4.a | 426.17 | 0 | 0 | 0 | 0 | 0 |
| 5.a | 44722 | 57 | 1251 | 1341 | 28 | 30 |
| 5.b | 6353.9 | 0 | 0 | 0 | 0 | 0 |
| 14.a | 10021.2 | 0 | 0 | 0 | 0 | 0 |
| Total | 721566 | 335 | 7241 | 31755 | 10 | 44 |

Table 4.4.1.5 Norwegian spring-spawning herring. Catch data provided by working group members and samples allocated to unsampled catches in SALLOC

| Countr <br> Y | Div | Q. | Catch <br> (T) | Samples allocated ('Fill in') |
| :---: | :---: | :---: | :---: | :---: |
| DE | 2a | 1 | 2.2 | NO_2a_q1,DK_2a_q1 |
| DE | 2a | 3 | 64.5 | IS_2a_q3,NL_2a_q3,RU_2a_q3 |
| DE | 2a | 4 | 5134.4 |  |
| DK | 2a | 1 | 14020.6 |  |
| DK | 2a | 4 | 5016.8 | ```NO_2a_q4,FO_2a_q4,IS_2a_q4,NL_2a_q4,RU_2a_q4,DE_2a_q 4``` |
| FO | 2a | 2 | 54.0 |  |
| FO | 2a | 3 | 7029.8 |  |
| FO | 2a | 4 | 84732.6 |  |
| FO | 5b | 2 | 125.2 | FO_2a_q2 |
| FO | 5b | 3 | 71.7 | FO_2a_q3 |
| FO | 5b | 4 | 6157.0 | FO_2a_q4 |
| GL | 14a | 2 | 1078.0 | RU_2a_q2 |
| GL | 14a | 3 | 8943.2 | IS_2a_q3,NL_2a_q3,RU_2a_q3,IS_5a_q3 |
| GL | 2a | 3 | 618.7 | IS_2a_q3,NL_2a_q3,RU_2a_q3 |
| GL | 2a | 4 | 1929.1 | $\begin{aligned} & \text { NO_2a_q4,FO_2a_q4,IS_2a_q4,NL_2a_q4,RU_2a_q4,DE_2a_q } \\ & 4 \end{aligned}$ |
| IR | 2a | 1 | 2315.8 |  |
| IR | 2a | 4 | 1178.9 |  |
| IS | 2a | 3 | 3358.0 |  |
| IS | 2a | 4 | 42320.0 |  |
| IS | 5a | 3 | 25446.0 |  |
| IS | 5a | 4 | 19276.0 |  |
| NL | 2a | 3 | 616.4 |  |
| NL | 2a | 4 | 5721.3 |  |
| NL | 4a | 4 | 341.2 | $\begin{aligned} & \text { NO_2a_q4,FO_2a_q4,IS_2a_q4,NL_2a_q4,RU_2a_q4,DE_2a_q } \\ & 4 \end{aligned}$ |
| NO | 2a | 1 | 144054.6 |  |
| NO | 2a | 2 | 2156.7 | NO_2a_q1 |
| NO | 2a | 3 | 773.2 | IS_2a_q3,NL_2a_q3,RU_2a_q3 |
| NO | 2a | 4 | 242313.9 |  |
| NO | 3a | 2 | 0.1 | NO_2a_q1 |
| NO | 4a | 1 | 56.7 | NO_2a_q1 |
| NO | 4a | 2 | 0.0 | NO_2a_q1 |
| NO | 4a | 3 | 28.3 | NO_2a_q4 |
| PL | 2a | 1 | 0.7 | NO_2a_q1,DK_2a_q1 |
| RU | 2a | 1 | 957.0 | NO_2a_q1,DK_2a_q1 |
| RU | 2a | 2 | 129.0 |  |
| RU | 2a | 3 | 9945.0 |  |
| RU | 2a | 4 | 80087.0 |  |
| SE | 2a | 1 | 405.0 | NO_2a_q1,DK_2a_q1 |
| SE | 2a | 4 | 750.0 | $\begin{aligned} & \text { NO_2a_q4,FO_2a_q4,IS_2a_q4,NL_2a_q4,RU_2a_q4,DE_2a_q } \\ & 4 \end{aligned}$ |


| Countr <br> $\mathbf{Y}$ | DIV <br> . | Q. | CATCH <br> $(\mathrm{T})$ | SAMPLES ALLOCATED ('fiLL IN') |
| :--- | :---: | :---: | :---: | :---: |
| UKS | 2a | 1 | 4356.2 | NO_2a_q1,DK_2a_q1 |
| UKS | 2 a | 4 | 1.7 | NO_2a_q4,FO_2a_q4,IS_2a_q4,NL_2a_q4,RU_2a_q4,DE_2a_q <br> 4 |

Table 4.4.3.1. Norwegian spring spawning herring. Catch in numbers (thousands).

|  |  |  |  |  |  |  |  | AGE |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1950 | 5112600 | 2000000 | 600000 | 276200 | 184800 | 185500 | 547000 | 628600 | 79500 | 88600 | 109500 | 86900 | 194500 | 368300 | 66400 | 344300 |
| 1951 | 1635500 | 7607700 | 400000 | 6600 | 383800 | 172400 | 164400 | 515600 | 602000 | 77100 | 82700 | 103100 | 107600 | 253500 | 348000 | 352500 |
| 1952 | 13721600 | 9149700 | 1232900 | 39300 | 60500 | 602300 | 136300 | 204500 | 380200 | 377900 | 79200 | 85700 | 107700 | 106800 | 186500 | 564400 |
| 1953 | 5697200 | 5055000 | 581300 | 740100 | 46600 | 100900 | 355600 | 81900 | 110900 | 314100 | 394900 | 61700 | 91200 | 94100 | 98800 | 730400 |
| 1954 | 10675990 | 7071090 | 855400 | 266300 | 1435500 | 142900 | 236000 | 490300 | 128100 | 199800 | 440400 | 460700 | 88400 | 100600 | 133000 | 803200 |
| 1955 | 5175600 | 2871100 | 510100 | 93000 | 276400 | 2045100 | 114300 | 189600 | 274700 | 85300 | 193400 | 295600 | 203200 | 58700 | 84600 | 580600 |
| 1956 | 5363900 | 2023700 | 627100 | 116500 | 251600 | 314200 | 2555100 | 110000 | 203900 | 264200 | 130700 | 198300 | 272800 | 163300 | 63000 | 565100 |
| 1957 | 5001900 | 3290800 | 219500 | 23300 | 373300 | 153800 | 228500 | 1985300 | 72000 | 127300 | 182500 | 88400 | 121200 | 149300 | 131600 | 281400 |
| 1958 | 9666990 | 2798100 | 666400 | 17500 | 17900 | 110900 | 89300 | 194400 | 973500 | 70700 | 123000 | 200900 | 98700 | 77400 | 70900 | 255600 |
| 1959 | 17896280 | 198530 | 325500 | 15100 | 26800 | 25900 | 146600 | 114800 | 240700 | 1103800 | 88600 | 124300 | 198000 | 88500 | 77400 | 235900 |
| 1960 | 12884310 | 13580790 | 392500 | 121700 | 18200 | 28100 | 24400 | 96200 | 73300 | 203900 | 1163000 | 85200 | 129700 | 153500 | 56700 | 168900 |
| 1961 | 6207500 | 16075600 | 2884800 | 31200 | 8100 | 4100 | 15000 | 19400 | 61600 | 49200 | 136100 | 728100 | 49700 | 45000 | 63000 | 60100 |
| 1962 | 3693200 | 4081100 | 1041300 | 1843800 | 8000 | 3100 | 7200 | 20200 | 11900 | 59100 | 52600 | 117000 | 813500 | 44200 | 54700 | 152300 |
| 1963 | 4807000 | 2119200 | 2045300 | 760400 | 835800 | 5300 | 1800 | 3600 | 18300 | 9300 | 107700 | 92500 | 174100 | 923700 | 79600 | 185300 |
| 1964 | 3613000 | 2728300 | 220300 | 114600 | 399000 | 2045800 | 13700 | 1500 | 3000 | 24900 | 29300 | 95600 | 82400 | 153000 | 772800 | 336800 |
| 1965 | 2303000 | 3780900 | 2853600 | 89900 | 256200 | 571100 | 2199700 | 19500 | 14900 | 7400 | 19100 | 40000 | 100500 | 107800 | 138700 | 883100 |
| 1966 | 3926500 | 662800 | 1678000 | 2048700 | 26900 | 466600 | 1306000 | 2884500 | 37900 | 14300 | 17400 | 26200 | 11000 | 69100 | 72100 | 556700 |
| 1967 | 426800 | 9877100 | 70400 | 1392300 | 3254000 | 26600 | 421300 | 1132000 | 1720800 | 8900 | 5700 | 3500 | 8500 | 8900 | 17500 | 104400 |
| 1968 | 1783600 | 437000 | 388300 | 99100 | 1880500 | 1387400 | 14220 | 94000 | 134100 | 345100 | 2000 | 1100 | 830 | 2500 | 2600 | 17000 |
| 1969 | 561200 | 507100 | 141900 | 188200 | 800 | 8800 | 4700 | 700 | 11700 | 33600 | 36000 | 300 | 200 | 200 | 200 | 2400 |
| 1970 | 119300 | 529400 | 33200 | 6300 | 18600 | 600 | 3300 | 3300 | 1000 | 13400 | 26200 | 28100 | 300 | 100 | 200 | 2000 |
| 1971 | 30500 | 42900 | 85100 | 1820 | 1020 | 1240 | 360 | 1110 | 1130 | 360 | 4410 | 6910 | 5450 | 0 | 20 | 120 |
| 1972 | 347100 | 41000 | 20400 | 35376 | 3476 | 3583 | 2481 | 694 | 1486 | 198 | 0 | 494 | 593 | 593 | 0 | 0 |
| 1973 | 29300 | 3500 | 1700 | 2389 | 25200 | 651 | 1506 | 278 | 178 | 0 | 0 | 0 | 0 | 0 | 180 | 0 |
| 1974 | 65900 | 7800 | 3900 | 100 | 241 | 24505 | 257 | 196 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  |  |  |  |  |  |  |  | AGE |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1975 | 30600 | 3600 | 1800 | 3268 | 132 | 910 | 30667 | 5 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | . 20100 | 2400 | 1200 | 23248 | 5436 | 0 | 0 | 13086 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 43000 | 6200 | 3100 | 22103 | 23595 | 336 | 0 | 419 | 10766 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 20100 | 2400 | 1200 | 3019 | 12164 | 20315 | 870 | 0 | 620 | 5027 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 32600 | 3800 | 1900 | 6352 | 1866 | 6865 | 11216 | 326 | 0 | 0 | 2534 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 6900 | 800 | 400 | 6407 | 5814 | 2278 | 8165 | 15838 | 441 | 8 | 0 | 2688 | 0 | 0 | 0 | 0 |
| 1981 | 8300 | 1100 | 11900 | 4166 | 4591 | 8596 | 2200 | 4512 | 8280 | 345 | 103 | 114 | 964 | 0 | 0 | 0 |
| 1982 | 22600 | 1100 | 200 | 13817 | 7892 | 4507 | 6258 | 1960 | 5075 | 6047 | 121 | 37 | 37 | 121 | 0 | 0 |
| 1983 | 127000 | 4680 | 1670 | 3183 | 21191 | 9521 | 6181 | 6823 | 1293 | 4598 | 7329 | 143 | 40 | 143 | 860 | 0 |
| 1984 | 33860 | 1700 | 2490 | 4483 | 5388 | 61543 | 18202 | 12638 | 15608 | 7215 | 16338 | 6478 | 0 | 0 | 0 | 1650 |
| 1985 | 28570 | 13150 | 207220 | 21500 | 15500 | 16500 | 130000 | 59000 | 55000 | 63000 | 10000 | 31000 | 50000 | 0 | 0 | 2640 |
| 1986 | 13810 | 1380 | 3090 | 539785 | 17594 | 14500 | 15500 | 105000 | 75000 | 42000 | 77000 | 19469 | 66000 | 80000 | 0 | 2470 |
| 1987 | 13850 | 6330 | 35770 | 19776 | 501393 | 18672 | 3502 | 7058 | 28000 | 12000 | 9500 | 4500 | 7834 | 6500 | 7000 | 450 |
| 1988 | 15490 | 2790 | 9110 | 62923 | 25059 | 550367 | 9452 | 3679 | 5964 | 14583 | 8872 | 2818 | 3356 | 2682 | 1560 | 540 |
| 1989 | 7120 | 1930 | 25200 | 2890 | 3623 | 5650 | 324290 | 3469 | 800 | 679 | 3297 | 1375 | 679 | 321 | 260 | 0 |
| 1990 | 1020 | 400 | 15540 | 18633 | 2658 | 11875 | 10854 | 226280 | 1289 | 1519 | 2036 | 2415 | 646 | 179 | 590 | 480 |
| 1991 | 100 | 3370 | 3330 | 8438 | 2780 | 1410 | 14698 | 8867 | 218851 | 2499 | 461 | 87 | 690 | 103 | 260 | 540 |
| 1992 | 1630 | 150 | 1340 | 12586 | 33100 | 4980 | 1193 | 11981 | 5748 | 225677 | 2483 | 639 | 247 | 1236 | 0 | 0 |
| 1993 | 6570 | 130 | 7240 | 28408 | 106866 | 87269 | 8625 | 3648 | 29603 | 18631 | 410110 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 430 | 20 | 8100 | 32500 | 110090 | 363920 | 164800 | 15580 | 8140 | 37330 | 35660 | 645410 | 2830 | 460 | 100 | 2070 |
| 1995 | 0 | 0 | 1130 | 57590 | 346460 | 622810 | 637840 | 231090 | 15510 | 15850 | 69750 | 83740 | 911880 | 4070 | 250 | 450 |
| 1996 | 0 | 0 | 30140 | 34360 | 713620 | 1571000 | 940580 | 406280 | 103410 | 5680 | 7370 | 66090 | 17570 | 836550 | 0 | 0 |
| 1997 | 0 | 0 | 21820 | 130450 | 270950 | 1795780 | 1993620 | 761210 | 326490 | 60870 | 20020 | 32400 | 90520 | 19120 | 370330 | 300 |
| 1998 | 0 | 0 | 82891 | 70323 | 242365 | 368310 | 1760319 | 1263750 | 381482 | 129971 | 42502 | 25343 | 3478 | 112604 | 5633 | 108514 |
| 1999 | 0 | 0 | 5029 | 137626 | 35820 | 134813 | 429433 | 1604959 | 1164263 | 291394 | 106005 | 14524 | 40040 | 7202 | 88598 | 63983 |
| 2000 | 0 | 0 | 14395 | 84016 | 560379 | 34933 | 110719 | 404460 | 1299253 | 1045001 | 216980 | 71589 | 16260 | 22701 | 23321 | 71811 |
| 2001 | 0 | 0 | 2076 | 102293 | 160678 | 426822 | 38749 | 95991 | 296460 | 839136 | 507106 | 73673 | 23722 | 3505 | 3356 | 22164 |
| 2002 | 0 | 0 | 62031 | 198360 | 643161 | 255516 | 326495 | 29843 | 93530 | 264675 | 663059 | 339326 | 52922 | 12437 | 7000 | 10087 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| 2003 | 0 | 3461 | 4524 | 75243 | 323958 | 730468 | 175878 | 167776 | 22866 | 74494 | 217108 | 567253 | 219097 | 38555 | 8111 | 6192 |
| 2004 | 125 | 1846 | 43800 | 24299 | 92300 | 429510 | 714433 | 111022 | 137940 | 26656 | 52467 | 169196 | 401564 | 210547 | 28028 | 11883 |
| 2005 | 0 | 442 | 20411 | 447788 | 94206 | 170547 | 643600 | 930309 | 121856 | 123291 | 37967 | 65289 | 139331 | 344822 | 126879 | 15697 |
| 2006 | 0 | 1968 | 45438 | 75824 | 729898 | 82107 | 171370 | 726041 | 772217 | 88701 | 77115 | 30339 | 57882 | 133665 | 142240 | 49128 |
| 2007 | 0 | 4475 | 8450 | 224636 | 366983 | 1804495 | 152916 | 242923 | 728836 | 511664 | 47215 | 25384 | 15316 | 24488 | 64755 | 58465 |
| 2008 | 0 | 39898 | 123949 | 36630 | 550274 | 670681 | 2295912 | 199592 | 256132 | 586583 | 369620 | 29633 | 36025 | 23775 | 25195 | 63176 |
| 2009 | 0 | 3468 | 113424 | 192641 | 149075 | 1193781 | 914748 | 1929631 | 142931 | 262037 | 423972 | 238174 | 45519 | 9337 | 10153 | 70538 |
| 2010 | 0 | 75981 | 61673 | 101948 | 209295 | 189784 | 1064866 | 711951 | 1421939 | 175010 | 180164 | 340781 | 179039 | 12558 | 11602 | 49773 |
| 2011 | 0 | 126972 | 249809 | 61706 | 104634 | 234330 | 210165 | 755382 | 543212 | 642787 | 90515 | 117230 | 136509 | 45082 | 6628 | 11638 |
| 2012 | 0 | 2680 | 13083 | 211630 | 49999 | 119627 | 281908 | 263330 | 747839 | 314694 | 357902 | 53109 | 44982 | 64273 | 12420 | 3604 |
| 2013 | 0 | 1 | 20715 | 60364 | 276901 | 71287 | 112558 | 283658 | 242243 | 591912 | 169525 | 145318 | 24936 | 10614 | 9725 | 2299 |
| 2014 | 0 | 265 | 1441 | 28301 | 57838 | 257529 | 50424 | 71721 | 194814 | 147083 | 381317 | 83050 | 57315 | 12746 | 1809 | 7501 |
| 2015 | 0 | 647 | 3244 | 16139 | 55749 | 52369 | 152347 | 34046 | 65728 | 156075 | 103393 | 201141 | 24310 | 49373 | 3369 | 6397 |
| 2016 | 0 | 197 | 2351 | 45483 | 43416 | 112147 | 85937 | 164454 | 52267 | 73576 | 174655 | 96476 | 179051 | 38546 | 32880 | 8379 |
| 2017 | 0 | 618 | 16390 | 64275 | 305483 | 114976 | 248192 | 162566 | 289931 | 98836 | 133145 | 276874 | 107473 | 220368 | 22357 | 49442 |

Table 4.4.4.1. Norwegian spring spawning herring. Weight at age in the catch (kg).

|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| 1950 | 0.007 | 0.025 | 0.058 | 0.110 | 0.188 | 0.211 | 0.234 | 0.253 | 0.266 | 0.280 | 0.294 | 0.303 | 0.312 | 0.32 | 0.323 | 0.334 |
| 1951 | 0.009 | 0.029 | 0.068 | 0.130 | 0.222 | 0.249 | 0.276 | 0.298 | 0.314 | 0.330 | 0.346 | 0.357 | 0.368 | 0.377 | 0.381 | 0.394 |
| 1952 | 0.008 | 0.026 | 0.061 | 0.115 | 0.197 | 0.221 | 0.245 | 0.265 | 0.279 | 0.293 | 0.308 | 0.317 | 0.327 | 0.335 | 0.339 | 0.349 |
| 1953 | 0.008 | 0.027 | 0.063 | 0.120 | 0.205 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.320 | 0.330 | 0.34 | 0.347 | 0.351 | 0.363 |
| 1954 | 0.008 | 0.026 | 0.062 | 0.117 | 0.201 | 0.225 | 0.250 | 0.269 | 0.284 | 0.299 | 0.313 | 0.323 | 0.333 | 0.341 | 0.345 | 0.356 |
| 1955 | 0.008 | 0.027 | 0.063 | 0.119 | 0.204 | 0.229 | 0.254 | 0.274 | 0.289 | 0.304 | 0.318 | 0.328 | 0.338 | 0.346 | 0.350 | 0.362 |
| 1956 | $0.008$ | 0.028 | 0.066 | 0.126 | 0.215 | 0.241 | 0.268 | 0.289 | 0.304 | 0.320 | 0.336 | 0.346 | 0.357 | 0.365 | 0.369 | 0.382 |
| 1957 | 0.008 | 0.028 | 0.066 | 0.127 | 0.216 | 0.243 | 0.269 | 0.290 | 0.306 | 0.322 | 0.338 | 0.348 | 0.359 | 0.367 | 0.371 | 0.384 |
| $1958$ | $0.009$ | $0.030$ | $0.070$ | 0.133 | 0.227 | 0.255 | 0.283 | 0.305 | 0.321 | 0.338 | 0.355 | 0.366 | 0.377 | 0.386 | 0.390 | 0.403 |
| 1959 | 0.009 | 0.030 | 0.071 | 0.135 | 0.231 | 0.259 | 0.287 | 0.310 | 0.327 | 0.344 | 0.360 | 0.372 | 0.383 | 0.392 | 0.397 | 0.409 |
| $1960$ | $0.006$ | 0.011 | 0.074 | 0.119 | 0.188 | 0.277 | 0.337 | 0.318 | 0.363 | 0.379 | 0.360 | 0.420 | 0.411 | 0.439 | 0.450 | 0.447 |
| $1961$ | $0.006$ | 0.010 | 0.045 | 0.087 | 0.159 | 0.276 | 0.322 | 0.372 | 0.363 | 0.393 | 0.407 | 0.397 | 0.422 | 0.447 | 0.465 | 0.452 |
| 1962 | $0.009$ | 0.023 | $0.055$ | $0.085$ | 0.148 | 0.288 | 0.333 | 0.360 | 0.352 | 0.350 | 0.374 | 0.384 | 0.374 | 0.394 | 0.399 | 0.414 |
| 1963 | 0.008 | 0.026 | 0.047 | 0.098 | 0.171 | 0.275 | 0.268 | 0.323 | 0.329 | 0.336 | 0.341 | 0.358 | 0.385 | 0.353 | 0.381 | 0.386 |
| 1964 | 0.009 | 0.024 | 0.059 | 0.139 | 0.219 | 0.239 | 0.298 | 0.295 | 0.339 | 0.350 | 0.358 | 0.351 | 0.367 | 0.375 | 0.372 | 0.433 |
| 1965 | $0.009$ | $0.016$ | $0.048$ | 0.089 | 0.217 | 0.234 | 0.262 | 0.331 | 0.360 | 0.367 | 0.386 | 0.395 | 0.393 | 0.404 | 0.401 | 0.431 |
| 1966 | 0.008 | 0.017 | 0.040 | 0.063 | 0.246 | 0.260 | 0.265 | 0.301 | 0.410 | 0.425 | 0.456 | 0.460 | 0.467 | 0.446 | 0.459 | 0.472 |
| 1967 | 0.009 | 0.015 | 0.036 | 0.066 | 0.093 | 0.305 | 0.305 | 0.310 | 0.333 | 0.359 | 0.413 | 0.446 | 0.401 | 0.408 | 0.439 | 0.430 |
| 1968 | $0.010$ | 0.027 | 0.049 | 0.075 | 0.108 | 0.158 | 0.375 | 0.383 | 0.364 | 0.382 | 0.441 | 0.410 |  | 0.517 | 0.491 | 0.485 |
| 1969 | 0.009 | 0.021 | 0.047 | 0.072 |  | 0.152 | 0.296 |  | 0.329 | 0.329 | 0.341 |  |  |  |  | 0.429 |
| 1970 | 0.008 | 0.058 | 0.085 | 0.105 | 0.171 |  | 0.216 | 0.277 | 0.298 | 0.304 | 0.305 | 0.309 |  |  |  | 0.376 |
| 1971 | 0.011 | 0.053 | 0.121 | 0.177 | 0.216 | 0.250 |  | 0.305 | 0.333 |  | 0.366 | 0.377 | 0.388 |  |  |  |
| 1972 | 0.011 | 0.029 | 0.062 | 0.103 | 0.154 | 0.215 | 0.258 |  | 0.322 |  |  |  |  |  |  |  |
| 1973 | 0.006 | 0.053 | 0.106 | 0.161 | 0.213 |  | 0.255 |  |  |  |  |  |  |  |  |  |
| 1974 | 0.006 | 0.055 | 0.117 |  |  | 0.249 |  |  |  |  |  |  |  |  |  |  |
| 1975 | 0.009 | 0.079 | 0.169 | 0.241 |  |  | 0.381 |  |  |  |  |  |  |  |  |  |


|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1976 | 0.007 | 0.062 | 0.132 | 0.189 | 0.250 |  |  | 0.323 |  |  |  |  |  |  |  |  |
| 1977 | 0.011 | 0.091 | 0.193 | 0.316 | 0.350 |  |  |  | 0.511 |  |  |  |  |  |  |  |
| 1978 | 0.012 | 0.100 | 0.210 | 0.274 | 0.424 | 0.454 |  |  |  | 0.613 |  |  |  |  |  |  |
| 1979 | 0.010 | 0.088 | 0.181 | 0.293 | 0.359 | 0.416 | 0.436 |  |  |  | 0.553 |  |  |  |  |  |
| 1980 | 0.012 |  |  | 0.266 | 0.399 | 0.449 | 0.460 | 0.485 |  |  |  | 0.608 |  |  |  |  |
| 1981 | 0.010 | 0.082 | 0.163 | 0.196 | 0.291 | 0.341 | 0.368 | 0.380 | 0.397 |  |  |  |  |  |  |  |
| 1982 | 0.010 | 0.087 | 0.159 | 0.256 | 0.312 | 0.378 | 0.415 | 0.435 | 0.449 | 0.448 |  |  |  |  |  |  |
| 1983 | 0.011 | 0.090 | 0.165 | 0.217 | 0.265 | 0.337 | 0.378 | 0.410 | 0.426 | 0.435 | 0.444 |  |  |  |  |  |
| 1984 | 0.009 | $0.047$ | 0.145 | 0.218 | 0.262 | 0.325 | $0.346$ | 0.381 | 0.400 | 0.413 | 0.405 | 0.426 |  |  |  | 0.415 |
| 1985 | 0.009 | 0.022 | 0.022 | 0.214 | 0.277 | 0.295 | 0.338 | 0.360 | 0.381 | 0.397 | 0.409 | 0.417 | 0.435 |  |  | 0.435 |
| 1986 | 0.007 | 0.077 | 0.097 | 0.055 | 0.249 | 0.294 | 0.312 | 0.352 | 0.374 | 0.398 | 0.402 | 0.401 | 0.410 | 0.410 |  | 0.410 |
| 1987 | 0.010 | 0.075 | 0.091 | 0.124 | 0.173 | 0.253 | 0.232 | 0.312 | 0.328 | 0.349 | 0.353 | 0.370 | 0.385 | 0.385 | 0.385 |  |
| 1988 | 0.008 | 0.062 | 0.075 | 0.124 | 0.154 | 0.194 | 0.241 | 0.265 | 0.304 | 0.305 | 0.317 | 0.308 | 0.334 | 0.334 | 0.334 |  |
| 1989 | 0.010 | 0.060 | 0.204 | 0.188 | 0.264 | 0.260 | 0.282 | 0.306 |  |  | 0.422 | 0.364 |  |  |  |  |
| 1990 | 0.007 |  | 0.102 | 0.230 | 0.239 | 0.266 | 0.305 | 0.308 | 0.376 | 0.407 | 0.412 | 0.424 |  |  |  |  |
| 1991 |  | 0.015 | 0.104 | 0.208 | 0.250 | 0.288 | 0.312 | 0.316 | 0.330 | 0.344 |  |  |  |  |  |  |
| 1992 | 0.007 |  | 0.103 | 0.191 | 0.233 | 0.304 | 0.337 | 0.365 | 0.361 | 0.371 | 0.403 |  |  | 0.404 |  |  |
| 1993 | 0.007 |  | 0.106 | 0.153 | 0.243 | 0.282 | 0.320 | 0.330 | 0.365 | 0.373 | 0.379 |  |  |  |  |  |
| 1994 |  |  | 0.102 | 0.194 | 0.239 | 0.280 | 0.317 | 0.328 | 0.356 | 0.372 | 0.390 | 0.379 | 0.399 | 0.403 |  |  |
| 1995 |  |  | 0.102 | 0.153 | 0.192 | 0.234 | 0.283 | 0.328 | 0.349 | 0.356 | 0.374 | 0.366 | 0.393 | 0.387 |  |  |
| 1996 |  |  | 0.136 | 0.136 | 0.168 | 0.206 | 0.262 | 0.309 | 0.337 | 0.366 | 0.360 | 0.361 | 0.367 | 0.379 |  |  |
| 1997 |  |  | 0.089 | 0.167 | 0.184 | 0.207 | 0.232 | 0.277 | 0.305 | 0.331 | 0.328 | 0.344 | 0.343 | 0.397 | 0.357 |  |
| 1998 |  |  | 0.111 | 0.150 | 0.216 | 0.221 | 0.249 | 0.277 | 0.316 | 0.338 | 0.374 | 0.372 | 0.366 | 0.396 | 0.377 | 0.406 |
| 1999 |  |  | 0.096 | 0.173 | 0.228 | 0.262 | 0.274 | 0.292 | 0.307 | 0.335 | 0.362 | 0.371 | 0.399 | 0.396 | 0.400 | 0.404 |
| 2000 |  |  | 0.124 | 0.175 | 0.222 | 0.242 | 0.289 | 0.303 | 0.310 | 0.328 | 0.349 | 0.383 | 0.411 | 0.410 | 0.419 | 0.409 |
| 2001 |  |  | 0.105 | 0.166 | 0.214 | 0.252 | 0.268 | 0.305 | 0.308 | 0.322 | 0.337 | 0.363 | 0.353 | 0.378 | 0.400 | 0.427 |
| 2002 |  |  | 0.056 | 0.128 | 0.198 | 0.255 | 0.281 | 0.303 | 0.322 | 0.323 | 0.334 | 0.345 | 0.369 | 0.407 | 0.410 | 0.435 |
| 2003 |  | 0.062 | 0.068 | 0.169 | 0.218 | 0.257 | 0.288 | 0.316 | 0.323 | 0.348 | 0.354 | 0.351 | 0.363 | 0.372 | 0.376 | 0.429 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 2004 | 0.022 | 0.066 | 0.143 | 0.18 | 0.227 | 0.26 | 0.29 | 0.323 | 0.355 | 0.375 | 0.383 | 0.399 | 0.395 | 0.405 | 0.429 | 0.439 |
| 2005 |  | 0.092 | 0.106 | 0.181 | 0.235 | 0.266 | 0.290 | 0.315 | 0.344 | 0.367 | 0.384 | 0.372 | 0.384 | 0.398 | 0.402 | 0.413 |
| 2006 |  | 0.055 | 0.102 | 0.171 | 0.238 | 0.268 | 0.292 | 0.311 | 0.330 | 0.365 | 0.374 | 0.376 | 0.388 | 0.396 | 0.398 | 0.407 |
| 2007 | 0.000 | 0.074 | 0.137 | 0.162 | 0.228 | 0.271 | 0.316 | 0.332 | 0.342 | 0.358 | 0.361 | 0.381 | 0.390 | 0.400 | 0.405 | 0.399 |
| 2008 | 0.000 | 0.026 | 0.106 | 0.145 | 0.209 | 0.254 | 0.296 | 0.318 | 0.341 | 0.353 | 0.363 | 0.367 | 0.395 | 0.396 | 0.386 | 0.413 |
| 2009 |  | 0.040 | 0.156 | 0.184 | 0.220 | 0.251 | 0.291 | 0.311 | 0.338 | 0.347 | 0.363 | 0.375 | 0.382 | 0.375 | 0.375 | 0.387 |
| 2010 |  | 0.059 | 0.107 | 0.177 | 0.218 | 0.261 | 0.279 | 0.311 | 0.325 | 0.343 | 0.362 | 0.370 | 0.388 | 0.391 | 0.376 | 0.441 |
| 2011 |  | 0.011 | 0.098 | 0.200 | 0.257 | 0.273 | 0.300 | 0.316 | 0.340 | 0.348 | 0.365 | 0.371 | 0.387 | 0.374 | 0.403 | 0.401 |
| 2012 |  | 0.034 | 0.126 | 0.211 | 0.272 | 0.301 | 0.308 | 0.331 | 0.335 | 0.351 | 0.354 | 0.370 | 0.389 | 0.389 | 0.382 | 0.388 |
| 2013 |  | 0.048 | 0.163 | 0.237 | 0.276 | 0.300 | 0.331 | 0.339 | 0.351 | 0.357 | 0.370 | 0.373 | 0.394 | 0.391 | 0.389 | 0.367 |
| 2014 |  | 0.057 | 0.179 | 0.233 | 0.271 | 0.293 | 0.322 | 0.342 | 0.353 | 0.367 | 0.365 | 0.374 | 0.375 | 0.378 | 0.418 | 0.371 |
| 2015 |  | 0.059 | 0.146 | 0.203 | 0.272 | 0.323 | 0.331 | 0.358 | 0.370 | 0.372 | 0.383 | 0.382 | 0.392 | 0.386 | 0.383 | 0.391 |
| 2016 |  | 0.048 | 0.111 | 0.212 | 0.255 | 0.290 | 0.333 | 0.339 | 0.361 | 0.367 | 0.370 | 0.381 | 0.378 | 0.388 | 0.383 | 0.395 |
| 2017 |  | 0.092 | 0.143 | 0.205 | 0.241 | 0.292 | 0.322 | 0.350 | 0.360 | 0.382 | 0.392 | 0.391 | 0.396 | 0.399 | 0.407 | 0.394 |

Table 4.4.4.2. Norwegian spring spawning herring. Weight at age in the stock (kg).

|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1950 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1951 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1952 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1953 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1954 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1955 | 0.001 | 0.008 | 0.047 | 0.100 | 0.195 | 0.213 | 0.260 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1956 | 0.001 | 0.008 | 0.047 | 0.100 | 0.205 | 0.230 | 0.249 | 0.275 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1957 | 0.001 | 0.008 | 0.047 | 0.100 | 0.136 | 0.228 | 0.255 | 0.262 | 0.290 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.364 |
| 1958 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.242 | 0.292 | 0.295 | 0.293 | 0.305 | 0.315 | 0.330 | 0.340 | 0.345 | 0.352 | 0.363 |
| 1959 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.252 | 0.260 | 0.290 | 0.300 | 0.305 | 0.315 | 0.325 | 0.330 | 0.340 | 0.345 | 0.358 |
| 1960 | 0.001 | 0.008 | 0.047 | 0.100 | 0.204 | 0.270 | 0.291 | 0.293 | 0.321 | 0.318 | 0.320 | 0.344 | 0.349 | 0.370 | 0.379 | 0.378 |
| 1961 | 0.001 | 0.008 | 0.047 | 0.100 | 0.232 | 0.250 | 0.292 | 0.302 | 0.304 | 0.323 | 0.322 | 0.321 | 0.344 | 0.357 | 0.363 | 0.368 |
| 1962 | 0.001 | 0.008 | 0.047 | 0.100 | 0.219 | 0.291 | 0.300 | 0.316 | 0.324 | 0.326 | 0.335 | 0.338 | 0.334 | 0.347 | 0.354 | 0.358 |
| 1963 | 0.001 | 0.008 | 0.047 | 0.100 | 0.185 | 0.253 | 0.294 | 0.312 | 0.329 | 0.327 | 0.334 | 0.341 | 0.349 | 0.341 | 0.358 | 0.375 |
| 1964 | 0.001 | 0.008 | 0.047 | 0.100 | 0.194 | 0.213 | 0.264 | 0.317 | 0.363 | 0.353 | 0.349 | 0.354 | 0.357 | 0.359 | 0.365 | 0.402 |
| 1965 | 0.001 | 0.008 | 0.047 | 0.100 | 0.186 | 0.199 | 0.236 | 0.260 | 0.363 | 0.350 | 0.370 | 0.360 | 0.378 | 0.387 | 0.390 | 0.394 |
| 1966 | 0.001 | 0.008 | 0.047 | 0.100 | 0.185 | 0.219 | 0.222 | 0.249 | 0.306 | 0.354 | 0.377 | 0.391 | 0.379 | 0.378 | 0.361 | 0.383 |
| 1967 | 0.001 | 0.008 | 0.047 | 0.100 | 0.180 | 0.228 | 0.269 | 0.270 | 0.294 | 0.324 | 0.420 | 0.430 | 0.366 | 0.368 | 0.433 | 0.414 |
| 1968 | 0.001 | 0.008 | 0.047 | 0.100 | 0.115 | 0.206 | 0.266 | 0.275 | 0.274 | 0.285 | 0.350 | 0.325 | 0.363 | 0.408 | 0.388 | 0.378 |
| 1969 | 0.001 | 0.008 | 0.047 | 0.100 | 0.115 | 0.145 | 0.270 | 0.300 | 0.306 | 0.308 | 0.318 | 0.340 | 0.368 | 0.360 | 0.393 | 0.397 |
| 1970 | 0.001 | 0.008 | 0.047 | 0.100 | 0.209 | 0.272 | 0.230 | 0.295 | 0.317 | 0.323 | 0.325 | 0.329 | 0.380 | 0.370 | 0.380 | 0.391 |
| 1971 | 0.001 | 0.015 | 0.080 | 0.100 | 0.190 | 0.225 | 0.250 | 0.275 | 0.290 | 0.310 | 0.325 | 0.335 | 0.345 | 0.355 | 0.365 | 0.390 |
| 1972 | 0.001 | 0.010 | 0.070 | 0.150 | 0.150 | 0.140 | 0.210 | 0.240 | 0.270 | 0.300 | 0.325 | 0.335 | 0.345 | 0.355 | 0.365 | 0.390 |
| 1973 | 0.001 | 0.010 | 0.085 | 0.170 | 0.259 | 0.342 | 0.384 | 0.409 | 0.404 | 0.461 | 0.520 | 0.534 | 0.500 | 0.500 | 0.500 | 0.500 |
| 1974 | 0.001 | 0.010 | 0.085 | 0.170 | 0.259 | 0.342 | 0.384 | 0.409 | 0.444 | 0.461 | 0.520 | 0.543 | 0.482 | 0.482 | 0.482 | 0.482 |
| 1975 | 0.001 | 0.010 | 0.085 | 0.181 | 0.259 | 0.342 | 0.384 | 0.409 | 0.444 | 0.461 | 0.520 | 0.543 | 0.482 | 0.482 | 0.482 | 0.482 |


|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1976 | 0.001 | 0.010 | 0.085 | 0.181 | 0.259 | 0.342 | 0.384 | 0.409 | 0.444 | 0.461 | 0.520 | 0.543 | 0.482 | 0.482 | 0.482 | 0.482 |
| 1977 | 0.001 | 0.010 | 0.085 | 0.181 | 0.259 | 0.343 | 0.384 | 0.409 | 0.444 | 0.461 | 0.520 | 0.543 | 0.482 | 0.482 | 0.482 | 0.482 |
| 1978 | 0.001 | 0.010 | 0.085 | 0.180 | 0.294 | 0.326 | 0.371 | 0.409 | 0.461 | 0.476 | 0.520 | 0.543 | 0.500 | 0.500 | 0.500 | 0.500 |
| 1979 | 0.001 | 0.010 | 0.085 | 0.178 | 0.232 | 0.359 | 0.385 | 0.420 | 0.444 | 0.505 | 0.520 | 0.551 | 0.500 | 0.500 | 0.500 | 0.500 |
| 1980 | 0.001 | 0.010 | 0.085 | 0.175 | 0.283 | 0.347 | 0.402 | 0.421 | 0.465 | 0.465 | 0.520 | 0.534 | 0.500 | 0.500 | 0.500 | 0.500 |
| 1981 | 0.001 | 0.010 | 0.085 | 0.170 | 0.224 | 0.336 | 0.378 | 0.387 | 0.408 | 0.397 | 0.520 | 0.543 | 0.512 | 0.512 | 0.512 | 0.512 |
| 1982 | 0.001 | 0.010 | 0.085 | 0.170 | 0.204 | 0.303 | 0.355 | 0.383 | 0.395 | 0.413 | 0.453 | 0.468 | 0.506 | 0.506 | 0.506 | 0.506 |
| 1983 | 0.001 | 0.010 | 0.085 | 0.155 | 0.249 | 0.304 | 0.368 | 0.404 | 0.424 | 0.437 | 0.436 | 0.493 | 0.495 | 0.495 | 0.495 | 0.495 |
| 1984 | 0.001 | 0.010 | 0.085 | 0.140 | 0.204 | 0.295 | 0.338 | 0.376 | 0.395 | 0.407 | 0.413 | 0.422 | 0.437 | 0.437 | 0.437 | 0.437 |
| 1985 | 0.001 | 0.010 | 0.085 | 0.148 | 0.234 | 0.265 | 0.312 | 0.346 | 0.370 | 0.395 | 0.397 | 0.428 | 0.428 | 0.428 | 0.428 | 0.428 |
| 1986 | 0.001 | 0.010 | 0.085 | 0.054 | 0.206 | 0.265 | 0.289 | 0.339 | 0.368 | 0.391 | 0.382 | 0.388 | 0.395 | 0.395 | 0.395 | 0.395 |
| 1987 | 0.001 | 0.010 | 0.055 | 0.090 | 0.143 | 0.241 | 0.279 | 0.299 | 0.316 | 0.342 | 0.343 | 0.362 | 0.376 | 0.376 | 0.376 | 0.376 |
| 1988 | 0.001 | 0.015 | 0.050 | 0.098 | 0.135 | 0.197 | 0.277 | 0.315 | 0.339 | 0.343 | 0.359 | 0.365 | 0.376 | 0.376 | 0.376 | 0.376 |
| 1989 | 0.001 | 0.015 | 0.100 | 0.154 | 0.175 | 0.209 | 0.252 | 0.305 | 0.367 | 0.377 | 0.359 | 0.395 | 0.396 | 0.396 | 0.396 | 0.396 |
| 1990 | 0.001 | 0.008 | 0.048 | 0.219 | 0.198 | 0.258 | 0.288 | 0.309 | 0.428 | 0.370 | 0.403 | 0.387 | 0.440 | 0.440 | 0.440 | 0.44 |
| 1991 | 0.001 | 0.011 | 0.037 | 0.147 | 0.210 | 0.244 | 0.300 | 0.324 | 0.336 | 0.343 | 0.382 | 0.366 | 0.425 | 0.425 | 0.425 | 0.425 |
| 1992 | 0.001 | 0.007 | 0.030 | 0.128 | 0.224 | 0.296 | 0.327 | 0.355 | 0.345 | 0.367 | 0.341 | 0.361 | 0.430 | 0.470 | 0.470 | 0.46 |
| 1993 | 0.001 | 0.008 | 0.025 | 0.081 | 0.201 | 0.265 | 0.323 | 0.354 | 0.358 | 0.381 | 0.369 | 0.396 | 0.393 | 0.374 | 0.403 | 0.4 |
| 1994 | 0.001 | 0.010 | 0.025 | 0.075 | 0.151 | 0.254 | 0.318 | 0.371 | 0.347 | 0.412 | 0.382 | 0.407 | 0.410 | 0.410 | 0.410 | 0.41 |
| 1995 | 0.001 | 0.018 | 0.025 | 0.066 | 0.138 | 0.230 | 0.296 | 0.346 | 0.388 | 0.363 | 0.409 | 0.414 | 0.422 | 0.410 | 0.410 | 0.426 |
| 1996 | 0.001 | 0.018 | 0.025 | 0.076 | 0.118 | 0.188 | 0.261 | 0.316 | 0.346 | 0.374 | 0.390 | 0.390 | 0.384 | 0.398 | 0.398 | 0.398 |
| 1997 | 0.001 | 0.018 | 0.025 | 0.096 | 0.118 | 0.174 | 0.229 | 0.286 | 0.323 | 0.370 | 0.378 | 0.386 | 0.360 | 0.393 | 0.391 | 0.391 |
| 1998 | 0.001 | 0.018 | 0.025 | 0.074 | 0.147 | 0.174 | 0.217 | 0.242 | 0.278 | 0.304 | 0.310 | 0.359 | 0.340 | 0.344 | 0.385 | 0.369 |
| 1999 | 0.001 | 0.018 | 0.025 | 0.102 | 0.150 | 0.223 | 0.240 | 0.264 | 0.283 | 0.315 | 0.345 | 0.386 | 0.386 | 0.386 | 0.382 | 0.395 |
| 2000 | 0.001 | 0.018 | 0.025 | 0.119 | 0.178 | 0.225 | 0.271 | 0.285 | 0.298 | 0.311 | 0.339 | 0.390 | 0.398 | 0.406 | 0.414 | 0.427 |
| 2001 | 0.001 | 0.018 | 0.025 | 0.075 | 0.178 | 0.238 | 0.247 | 0.296 | 0.307 | 0.314 | 0.328 | 0.351 | 0.376 | 0.406 | 0.414 | 0.425 |
| 2002 | 0.001 | 0.010 | 0.023 | 0.057 | 0.177 | 0.241 | 0.275 | 0.302 | 0.311 | 0.314 | 0.328 | 0.341 | 0.372 | 0.405 | 0.415 | 0.438 |
| 2003 | 0.001 | 0.010 | 0.055 | 0.098 | 0.159 | 0.211 | 0.272 | 0.305 | 0.292 | 0.331 | 0.337 | 0.347 | 0.356 | 0.381 | 0.414 | 0.433 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 2004 | 0.001 | 0.010 | 0.055 | 0.106 | 0.149 | 0.212 | 0.241 | 0.279 | 0.302 | 0.337 | 0.354 | 0.355 | 0.360 | 0.371 | 0.400 | 0.429 |
| 2005 | 0.001 | 0.010 | 0.046 | 0.112 | 0.156 | 0.234 | 0.267 | 0.295 | 0.330 | 0.363 | 0.377 | 0.414 | 0.406 | 0.308 | 0.420 | 0.452 |
| 2006 | 0.001 | 0.010 | 0.042 | 0.107 | 0.179 | 0.232 | 0.272 | 0.297 | 0.318 | 0.371 | 0.365 | 0.393 | 0.395 | 0.399 | 0.415 | 0.428 |
| 2007 | 0.001 | 0.010 | 0.036 | 0.086 | 0.155 | 0.226 | 0.265 | 0.312 | 0.310 | 0.364 | 0.384 | 0.352 | 0.386 | 0.304 | 0.420 | 0.412 |
| 2008** | 0.001 | 0.010 | 0.044 | 0.077 | 0.146 | 0.212 | 0.269 | 0.289 | 0.327 | 0.351 | 0.358 | 0.372 | 0.411 | 0.353 | 0.389 | 0.393 |
| 2009*** | 0.001 | 0.010 | 0.044 | 0.077 | 0.141 | 0.215 | 0.270 | 0.306 | 0.336 | 0.346 | 0.364 | 0.369 | 0.411 | 0.353 | 0.389 | 0.393 |
| 2010**** | 0.001 | 0.01 | 0.044 | 0.077 | 0.188 | 0.22 | 0.251 | 0.286 | 0.308 | 0.333 | 0.344 | 0.354 | 0.373 | 0.353 | 0.389 | 0.393 |
| 2011 | 0.001 | 0.01 | 0.044 | 0.118 | 0.185 | 0.209 | 0.246 | 0.277 | 0.310 | 0.322 | 0.339 | 0.349 | 0.364 | 0.363 | 0.389 | 0.393 |
| 2012 | 0.001 | 0.01 | 0.044 | 0.138 | 0.185 | 0.256 | 0.273 | 0.290 | 0.305 | 0.330 | 0.342 | 0.361 | 0.390 | 0.377 | 0.389 | 0.393 |
| 2013 | 0.001 | 0.01 | 0.044 | 0.138 | 0.204 | 0.267 | 0.305 | 0.309 | 0.320 | 0.328 | 0.346 | 0.350 | 0.390 | 0.377 | 0.389 | 0.393 |
| 2014 | 0.001 | 0.01 | 0.044 | 0.138 | 0.198 | 0.274 | 0.301 | 0.326 | 0.333 | 0.339 | 0.347 | 0.344 | 0.362 | 0.362 | 0.389 | 0.393 |
| 2015 | 0.001 | 0.01 | 0.044 | 0.138 | 0.187 | 0.243 | 0.299 | 0.326 | 0.319 | 0.345 | 0.346 | 0.354 | 0.382 | 0.376 | 0.389 | 0.393 |
| 2016 | 0.001 | 0.01 | 0.054 | 0.115 | 0.186 | 0.247 | 0.293 | 0.320 | 0.334 | 0.353 | 0.354 | 0.352 | 0.361 | 0.370 | 0.380 | 0.388 |
| 2017 | 0.001 | 0.01 | 0.054 | 0.115 | 0.190 | 0.247 | 0.282 | 0.322 | 0.338 | 0.351 | 0.359 | 0.361 | 0.361 | 0.368 | 0.380 | 0.386 |
| 2018 | 0.001 | 0.01 | 0.054 | 0.115 | 0.149 | 0.225 | 0.260 | 0.289 | 0.312 | 0.343 | 0.359 | 0.361 | 0.369 | 0.368 | 0.377 | 0.386 |

** mean weight at ages 11 and 13 are mean of 5 previous years at the same age. These age groups were not present in the catches of the wintering survey from which the stock weight are derived.
*** derived from catch data from the wintering area north of $69^{\circ} \mathrm{N}$ during December 2008 - January 2009 for age groups 4-11.
${ }^{* * * *}$ derived from catch data from the wintering area north of $69^{\circ} \mathrm{N}$ during January 2010 for age groups 4-12.

Table 4.4.5.1. Norwegian Spring-spawning herring. Mature at age. The time-series was provided by WKHERMAT in 2010 and are used in the assessment since 2010.

| Year/Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1951 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1952 | 0 | 0 | 0 | 0 | 0.1 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1953 | 0 | 0 | 0 | 0 | 0.3 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1954 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1955 | 0 | 0 | 0 | 0.1 | 0.4 | 0.4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1956 | 0 | 0 | 0 | 0 | 0.5 | 0.7 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1957 | 0 | 0 | 0 | 0 | 0.3 | 0.8 | 0.8 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1958 | 0 | 0 | 0 | 0 | 0.3 | 0.5 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1959 | 0 | 0 | 0 | 0 | 0.7 | 0.8 | 1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1960 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1961 | 0 | 0 | 0 | 0 | 0.1 | 0.8 | 1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1962 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1963 | 0 | 0 | 0 | 0 | 0.1 | 0.4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1964 | 0 | 0 | 0 | 0 | 0.1 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1965 | 0 | 0 | 0 | 0 | 0.5 | 0.4 | 0.9 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1966 | 0 | 0 | 0 | 0 | 0.5 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1967 | 0 | 0 | 0 | 0 | 0.3 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1968 | 0 | 0 | 0 | 0 | 0 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1969 | 0 | 0 | 0 | 0.1 | 0.2 | 0.3 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1970 | 0 | 0 | 0 | 0 | 0.4 | 0.3 | 0.4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1971 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1972 | 0 | 0 | 0 | 0 | 0.4 | 0.3 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1973 | 0 | 0 | 0 | 0.1 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1974 | 0 | 0 | 0 | 0 | 0.6 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1975 | 0 | 0 | 0 | 0.1 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1976 | 0 | 0 | 0 | 0.1 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1977 | 0 | 0 | 0 | 0.3 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1978 | 0 | 0 | 0 | 0.2 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1979 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1980 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1981 | 0 | 0 | 0 | 0.1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1982 | 0 | 0 | 0 | 0.1 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1983 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1984 | 0 | 0 | 0 | 0.1 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1985 | 0 | 0 | 0 | 0.1 | 0.8 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0 | 0 | 0 | 0 | 0.5 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0 | 0 | 0 | 0.1 | 0.8 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0 | 0 | 0 | 0 | 0.2 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0 | 0 | 0 | 0.2 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0 | 0 | 0 | 0 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0 | 0 | 0 | 0 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1993 | 0 | 0 | 0 | 0 | 0.5 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |


| Year/AGE | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5 +}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 0 | 0 | 0 | 0 | 0.1 | 0.9 | $\mathbf{1}$ | $\mathbf{1}$ | $\mathbf{1}$ | $\mathbf{1}$ | $\mathbf{1}$ | 1 | 1 | 1 | 1 | 1 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1997 | 0 | 0 | 0 | 0.1 | 0 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1998 | 0 | 0 | 0 | 0 | 0.6 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1999 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2000 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2001 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2002 | 0 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2003 | 0 | 0 | 0 | 0 | 0.2 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2004 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2005 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2006 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2007 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2008 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2009 | 0 | 0 | 0 | 0 | 0.1 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2010 | 0 | 0 | 0 | 0 | 0.2 | 0.4 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2011 | 0 | 0 | 0 | 0 | 0.4 | 0.7 | 0.8 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2012 | 0 | 0 | 0 | 0 | 0.5 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2013 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2014 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2015 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2016 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2017 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2018 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 4.4.7.1. Norwegian Spring-spawning herring. Estimated indices (with StoX) from the acoustic surveys on the spawning grounds in February-March. Numbers in millions. Biomass in thousand tonnes. "Fleet 1"

| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total | BIomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0 | 375 | 299 | 8066 | 86 | 33 | 11 | 38 | 22 | 41 | 0 | 0 | 0 | 0 | 8970 | 1631 |
| 1989 | 164 | 17 | 336 | 89 | 3995 | 106 | 12 | 8 | 59 | 0 | 4 | 39 | 0 | 8 | 4835 | 1175 |
| 1990 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 1991 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 1992* | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 1993* | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 1994 | 43 | 99 | 48 | 851 | 480 | 73 | 15 | 152 | 43 | 1838 | 3 | 3 | 0 | 0 | 3651 | 1215 |
| 1995 | 4 | 409 | 4643 | 3186 | 1986 | 292 | 18 | 0 | 141 | 76 | 2299 | 0 | 0 | 0 | 13053 | 3669 |
| 1996 | 126 | 147 | 1885 | 7923 | 2384 | 887 | 314 | 0 | 0 | 121 | 0 | 1830 | 0 | 0 | 15616 | 3382 |
| 1997* | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 1998 | 41 | 330 | 984 | 3012 | 13089 | 8214 | 1909 | 588 | 194 | 35 | 0 | 359 | 0 | 1415 | 30169 | 7008 |
| 1999 | 119 | 1572 | 379 | 1366 | 2593 | 9356 | 6979 | 1632 | 495 | 124 | 0 | 0 | 360 | 359 | 25333 | 6235 |
| 2000 | 1399 | 672 | 2617 | 103 | 485 | 1139 | 4193 | 2864 | 547 | 48 | 2 | 0 | 15 | 217 | 14301 | 3282 |
| 2001** | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 2002** | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 2003** | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 2004** | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 2005 | 39 | 270 | 662 | 2086 | 5871 | 8223 | 660 | 457 | 183 | 113 | 557 | 1138 | 595 | 6 | 20859 | 5223 |
| 2006 | 27 | 98 | 6073 | 478 | 912 | 3291 | 3290 | 122 | 67 | 25 | 72 | 54 | 265 | 63 | 14836 | 3392 |
| 2007 | 32 | 369 | 1594 | 12175 | 622 | 646 | 2842 | 3258 | 137 | 223 | 34 | 179 | 262 | 554 | 22925 | 5238 |
| 2008 | 15 | 70 | 2449 | 2699 | 9060 | 530 | 476 | 1599 | 1600 | 153 | 104 | 49 | 138 | 152 | 19094 | 4581 |
| 2009 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 2010 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 2011 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 2012 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 2013 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |


| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total | Biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |  |
| 2015 | 230 | 516 | 2748 | 768 | 3223 | 377 | 650 | 2868 | 720 | 7251 | 336 | 1733 | 50 | 229 | 21712 | 6390 |
| 2016 | 17 | 218 | 253 | 539 | 404 | 2288 | 242 | 569 | 2792 | 681 | 4144 | 197 | 982 | 107 | 13433 | 4338 |
| 2017 | 13 | 95 | 1078 | 666 | 868 | 411 | 1376 | 176 | 231 | 1903 | 295 | 2600 | 74 | 697 | 10486 | 3295 |
| 2018 | 95 | 145 | 1779 | 2780 | 485 | 824 | 622 | 1083 | 463 | 378 | 1188 | 360 | 1524 | 321 | 12047 | 3260 |

Table 4.4.7.2. Norwegian spring-spawning herring. Acoustic estimates (billion individuals) of immature herring in the Barents Sea in May/June from IESNS. Values in the years 2009-2017 are estimated with StoX. "Fleet 4"

|  | AGE |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 |
| 1991 | 24.3 | 5.2 |  |  |  |
| 1992 | 32.6 | 14 | 5.7 |  |  |
| 1993 | 102.7 | 25.8 | 1.5 |  |  |
| 1994 | 6.6 | 59.2 | 18 | 1.7 |  |
| 1995 | 0.5 | 7.7 | 8 | 1.1 |  |
| 1996* | 0.1 | 0.25 | 1.8 | 0.6 | 0.03 |
| 1997** | 2.6 | 0.04 | 0.4 | 0.35 | 0.05 |
| 1998 | 9.5 | 4.7 | 0.01 | 0.01 | 0 |
| 1999 | 49.5 | 4.9 | 0 | 0 | 0 |
| 2000 | 105.4 | 27.9 | 0 | 0 | 0 |
| 2001 | 0.3 | 7.6 | 8.8 | 0 | 0 |
| $2002$ | 0.5 | 3.9 | 0 | 0 | 0 |
| 2003*** |  |  |  |  |  |
| $2004 * * *$ |  |  |  |  |  |
| 2005 | 23.3 | 4.5 | 2.5 | 0.4 | 0.3 |
| 2006 | 3.7 | 35.0 | 5.3 | 0.87 | 0 |
| 2007 | 2.1 | 3.7 | 12.5 | 1.9 | 0 |
| $2008^{\wedge}$ |  |  |  |  |  |
| 2009 | 0.286 | 0.286 | 0.215 | 0.072 | 0 |
| 2010 | 5.121 | 1.366 | 0 | 0 | 0 |
| 2011 | 1.079 | 3.802 | 0.039 | 0 | 0 |
| 2012 | 0.884 | 0.015 | 0 | 0 | 0 |
| 2013 | 0.132 | 1.982 | 0.264 | 0.088 | 0 |
| 2014 | 3.727 | 3.055 | 1.797 | 0.131 | 0.044 |
| 2015 | 0.33 | 11.471 | 1.218 | 0.198 | 0 |
| 2016 | 1.677 | 5.463 | 1.668 | 0.103 | 0.042 |
| 2017 | 14.658 | 3.266 | 0 | 0 | 0 |
| 2018 | 6.866 | 17.404 | 0.943 | 0.009 | 0 |

*Average of Norwegian and Russian estimates
**Combination of Norwegian and Russian estimates as described in 1998 WG report, since then only Russian estimates
***No surveys
${ }^{\wedge}$ Not a full survey

Table 4.4.7.3. Norwegian spring-spawning herring. Estimates from the international acoustic survey on the feeding areas in the Norwegian Sea in May (IESNS). Numbers in millions. Biomass in thousands. Values in the years 2008-2017 are estimated indices by StoX. "Fleet 5"

|  |  |  |  |  |  |  |  | Age |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ | Total | Biomass |
| 1996 | 0 | 0 | 4114 | 22461 | 13244 | 4916 | 2045 | 424 | 14 | 7 | 155 | 0 | 3134 |  |  | 50514 | 8532 |
| 1997 | 0 | 0 | 1169 | 3599 | 18867 | 13546 | 2473 | 1771 | 178 | 77 | 288 | 190 | 60 | 2697 |  | 44915 | 9435 |
| 1998 | 24 | 1404 | 367 | 1099 | 4410 | 16378 | 10160 | 2059 | 804 | 183 | 0 | 0 | 35 | 0 | 492 | 37415 | 8004 |
| 1999 | 0 | 215 | 2191 | 322 | 965 | 3067 | 11763 | 6077 | 853 | 258 | 5 | 14 | 0 | 158 | 128 | 26016 | 6299 |
| 2000 | 0 | 157 | 1353 | 2783 | 92 | 384 | 1302 | 7194 | 5344 | 1689 | 271 | 0 | 114 | 0 | 75 | 20758 | 6001 |
| 2001 | 0 | 1540 | 8312 | 1430 | 1463 | 179 | 204 | 3215 | 5433 | 1220 | 94 | 178 | 0 | 0 | 6 | 23274 | 3937 |
| 2002 | 0 | 677 | 6343 | 9619 | 1418 | 779 | 375 | 847 | 1941 | 2500 | 1423 | 61 | 78 | 28 | 0 | 26089 | 4628 |
| 2003 | 32073 | 8115 | 6561 | 9985 | 9961 | 1499 | 732 | 146 | 228 | 1865 | 2359 | 1769 |  | 287 | 0 | 75580 | 6653 |
| 2004 | 0 | 13735 | 1543 | 5227 | 12571 | 10710 | 1075 | 580 | 76 | 313 | 362 | 1294 | 1120 | 10 | 88 | 48704 | 7687 |
| 2005 | 0 | 1293 | 19679 | 1353 | 1765 | 6205 | 5371 | 651 | 388 | 139 | 262 | 526 | 1003 | 364 | 115 | 39114 | 5109 |
| 2006 | 0 | 19 | 306 | 14560 | 1396 | 2011 | 6521 | 6978 | 679 | 713 | 173 | 407 | 921 | 618 | 243 | 35545 | 9100 |
| 2007 | 0 | 411 | 2889 | 5877 | 20292 | 1260 | 1992 | 6780 | 5582 | 647 | 488 | 372 | 403 | 1048 | 1010 | 49051 | 12161 |
| 2008 | 0 | 1240 | 631 | 10809 | 8271 | 14827 | 1513 | 2257 | 4848 | 2734 | 449 | 149 | 151 | 270 | 491 | 48665 | 10558 |
| 2009 | 0 | 144 | 1669 | 2159 | 12300 | 8994 | 9527 | 2147 | 1435 | 2466 | 1411 | 188 | 193 | 123 | 231 | 43082 | 9728 |
| 2010 | 234 | 125 | 542 | 2334 | 1781 | 8351 | 5988 | 5601 | 869 | 882 | 983 | 578 | 90 | 72 | 57 | 28622 | 6633 |
| 2011 | 0 | 1205 | 977 | 1528 | 3607 | 2564 | 9420 | 4542 | 4298 | 825 | 892 | 712 | 261 | 37 | 39 | 30917 | 7395 |
| 2012 | 0 | 378 | 2895 | 412 | 670 | 1646 | 2560 | 4226 | 2026 | 2097 | 298 | 607 | 315 | 155 | 47 | 18331 | 4435 |
| 2013 | 0 | 205 | 776 | 3955 | 434 | 1211 | 2036 | 3070 | 4652 | 2767 | 1873 | 692 | 805 | 186 | 83 | 22747 | 5888 |
| 2014 | 17 | 517 | 1231 | 798 | 2790 | 749 | 1065 | 2681 | 2285 | 2842 | 1119 | 778 | 350 | 76 | 198 | 17505 | 4555 |
| 2015 | 0 | 385 | 468 | 1299 | 1176 | 3548 | 1399 | 1160 | 3178 | 2523 | 4350 | 712 | 788 | 262 | 194 | 21443 | 5846 |
| 2016 | 0 | 75 | 3549 | 1508 | 2215 | 1779 | 2683 | 929 | 1143 | 1770 | 1851 | 2877 | 928 | 439 | 136 | 21889 | 5419 |
| 2017 | 11 | 132 | 1063 | 4363 | 1192 | 1522 | 874 | 1453 | 327 | 727 | 975 | 1785 | 2229 | 538 | 238 | 17441 | 4203 |
| 2018 | 0 | 500 | 1052 | 2063 | 5686 | 973 | 1434 | 561 | 1328 | 338 | 689 | 1565 | 1478 | 1529 | 488 | 19684 | 5042 |

Table 4.4.8.1 Norwegian spring-spawning herring. Relative standard error of estimated catch-at-age used by XSAM.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.346 | 0.205 | 0.263 | 0.114 | 0.343 | 0.442 | 0.388 | 0.305 | 0.349 | 0.475 | 0.357 |
| 1989 | 0.263 | 0.472 | 0.444 | 0.394 | 0.132 | 0.449 | 0.668 | 0.698 | 0.455 | 0.577 | 0.591 |
| 1990 | 0.3 | 0.285 | 0.483 | 0.322 | 0.33 | 0.145 | 0.587 | 0.562 | 0.519 | 0.495 | 0.529 |
| 1991 | 0.454 | 0.353 | 0.477 | 0.573 | 0.304 | 0.349 | 0.147 | 0.491 | 0.775 | 1.216 | 0.554 |
| 1992 | 0.581 | 0.317 | 0.244 | 0.407 | 0.599 | 0.321 | 0.392 | 0.145 | 0.492 | 0.71 | 0.565 |
| 1993 | 0.368 | 0.255 | 0.178 | 0.188 | 0.351 | 0.443 | 0.252 | 0.285 | 0.124 | NA | NA |
| 1994 | 0.357 | 0.246 | 0.177 | 0.128 | 0.158 | 0.299 | 0.357 | 0.236 | 0.239 | 0.11 | 0.397 |
| 1995 | 0.608 | 0.21 | 0.13 | 0.111 | 0.11 | 0.145 | 0.3 | 0.298 | 0.2 | 0.19 | 0.1 |
| 1996 | 0.251 | 0.242 | 0.107 | 0.086 | 0.099 | 0.124 | 0.18 | 0.393 | 0.367 | 0.203 | 0.102 |
| 1997 | 0.273 | 0.169 | 0.138 | 0.083 | 0.081 | 0.105 | 0.132 | 0.207 | 0.28 | 0.246 | 0.119 |
| 1998 | 0.191 | 0.199 | 0.143 | 0.127 | 0.084 | 0.091 | 0.126 | 0.169 | 0.228 | 0.263 | 0.145 |
| 1999 | 0.406 | 0.166 | 0.239 | 0.167 | 0.122 | 0.086 | 0.093 | 0.136 | 0.178 | 0.305 | 0.15 |
| 2000 | 0.306 | 0.19 | 0.114 | 0.241 | 0.176 | 0.124 | 0.091 | 0.096 | 0.147 | 0.198 | 0.167 |
| 2001 | 0.516 | 0.18 | 0.159 | 0.122 | 0.234 | 0.183 | 0.135 | 0.102 | 0.117 | 0.197 | 0.215 |
| 2002 | 0.206 | 0.151 | 0.11 | 0.141 | 0.132 | 0.251 | 0.185 | 0.139 | 0.109 | 0.13 | 0.191 |
| 2003 | 0.418 | 0.196 | 0.132 | 0.106 | 0.156 | 0.158 | 0.27 | 0.196 | 0.147 | 0.113 | 0.138 |
| 2004 | 0.227 | 0.266 | 0.185 | 0.122 | 0.107 | 0.176 | 0.166 | 0.259 | 0.216 | 0.157 | 0.109 |
| 2005 | 0.278 | 0.121 | 0.184 | 0.157 | 0.11 | 0.099 | 0.172 | 0.171 | 0.235 | 0.203 | 0.11 |
| 2006 | 0.224 | 0.195 | 0.106 | 0.191 | 0.157 | 0.106 | 0.104 | 0.187 | 0.194 | 0.25 | 0.126 |
| 2007 | 0.353 | 0.146 | 0.128 | 0.083 | 0.162 | 0.143 | 0.106 | 0.117 | 0.222 | 0.262 | 0.159 |
| 2008 | 0.171 | 0.238 | 0.114 | 0.108 | 0.078 | 0.15 | 0.141 | 0.112 | 0.127 | 0.252 | 0.163 |
| 2009 | 0.175 | 0.152 | 0.163 | 0.093 | 0.1 | 0.081 | 0.165 | 0.14 | 0.123 | 0.143 | 0.167 |
| 2010 | 0.207 | 0.18 | 0.148 | 0.152 | 0.096 | 0.107 | 0.088 | 0.156 | 0.155 | 0.13 | 0.141 |
| 2011 | 0.142 | 0.206 | 0.179 | 0.144 | 0.148 | 0.105 | 0.115 | 0.11 | 0.186 | 0.174 | 0.15 |
| 2012 | 0.314 | 0.148 | 0.219 | 0.173 | 0.137 | 0.14 | 0.105 | 0.133 | 0.128 | 0.215 | 0.171 |
| 2013 | 0.277 | 0.208 | 0.138 | 0.199 | 0.176 | 0.137 | 0.143 | 0.112 | 0.157 | 0.164 | 0.222 |
| 2014 | 0.57 | 0.255 | 0.21 | 0.14 | 0.218 | 0.198 | 0.151 | 0.163 | 0.126 | 0.191 | 0.193 |
| 2015 | 0.458 | 0.297 | 0.212 | 0.216 | 0.162 | 0.242 | 0.203 | 0.161 | 0.18 | 0.15 | 0.19 |
| 2016 | 0.499 | 0.224 | 0.227 | 0.176 | 0.189 | 0.158 | 0.216 | 0.197 | 0.156 | 0.183 | 0.14 |
| 2017 | 0.295 | 0.204 | 0.134 | 0.175 | 0.142 | 0.159 | 0.136 | 0.182 | 0.168 | 0.138 | 0.125 |
| 2018 | 0.331 | 0.222 | 0.194 | 0.179 | 0.178 | 0.189 | 0.207 | 0.218 | 0.232 | 0.291 | 0.247 |

Table 4.4.8.2 Norwegian spring-spawning herring. Relative standard error of Fleet 1 used by XSAM.

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.334 | 0.353 | 0.159 | 0.476 | 0.599 | 0.781 | 0.579 | 0.661 | 0.569 | NA |
| 1989 | 0.703 | 0.343 | 0.472 | 0.189 | 0.453 | 0.765 | 0.843 | 0.521 | NA | 0.54 |
| 1990 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1991 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1992 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1993 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1994 | 0.46 | 0.548 | 0.274 | 0.315 | 0.495 | 0.725 | 0.415 | 0.562 | 0.228 | 0.904 |
| 1995 | 0.327 | 0.182 | 0.199 | 0.223 | 0.355 | 0.694 | NA | 0.422 | 0.49 | 0.216 |
| 1996 | 0.418 | 0.226 | 0.16 | 0.214 | 0.271 | 0.348 | NA | NA | 0.438 | 0.228 |
| 1997 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1998 | 0.344 | 0.265 | 0.202 | 0.142 | 0.159 | 0.226 | 0.3 | 0.391 | 0.591 | 0.23 |
| 1999 | 0.236 | 0.333 | 0.244 | 0.21 | 0.154 | 0.165 | 0.234 | 0.312 | 0.436 | 0.285 |
| 2000 | 0.29 | 0.209 | 0.456 | 0.314 | 0.255 | 0.187 | 0.205 | 0.305 | 0.548 | 0.374 |
| 2001 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2002 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2003 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2004 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2005 | 0.361 | 0.291 | 0.221 | 0.172 | 0.159 | 0.291 | 0.318 | 0.397 | 0.446 | 0.216 |
| 2006 | 0.461 | 0.171 | 0.315 | 0.269 | 0.198 | 0.198 | 0.437 | 0.505 | 0.641 | 0.319 |
| 2007 | 0.335 | 0.236 | 0.144 | 0.296 | 0.293 | 0.205 | 0.198 | 0.425 | 0.378 | 0.262 |
| 2008 | 0.5 | 0.212 | 0.208 | 0.155 | 0.307 | 0.315 | 0.235 | 0.235 | 0.414 | 0.321 |
| 2009 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2010 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2011 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2012 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2013 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2014 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2015 | 0.309 | 0.207 | 0.281 | 0.199 | 0.333 | 0.292 | 0.205 | 0.285 | 0.164 | 0.215 |
| 2016 | 0.38 | 0.367 | 0.306 | 0.328 | 0.216 | 0.371 | 0.302 | 0.206 | 0.289 | 0.175 |
| 2017 | 0.465 | 0.259 | 0.291 | 0.273 | 0.327 | 0.244 | 0.4 | 0.375 | 0.226 | 0.193 |
| 2018 | 0.42 | 0.229 | 0.206 | 0.314 | 0.276 | 0.296 | 0.259 | 0.317 | 0.333 | 0.196 |

Table 4.4.8.3 Norwegian spring-spawning herring. Relative standard error of Fleet 4 used by XSAM.

| Year/AGE | 2 |
| :---: | :---: |
| 1991 | 0.351 |
| 1992 | 0.337 |
| 1993 | 0.286 |
| 1994 | 0.423 |
| 1995 | 0.61 |
| 1996 | 0.767 |
| 1997 | 0.483 |
| 1998 | 0.402 |
| 1999 | 0.318 |
| 2000 | 0.285 |
| 2001 | 0.656 |
| 2002 | 0.61 |
| 2003 | NA |
| 2004 | NA |
| 2005 | 0.354 |
| 2006 | 0.459 |
| 2007 | 0.498 |
| 2008 | 0.865 |
| 2009 | 0.661 |
| 2010 | 0.439 |
| 2011 | 0.547 |
| 2012 | 0.563 |
| 2013 | 0.738 |
| 2014 | 0.459 |
| 2015 | 0.648 |
| 2016 | 0.514 |
| 2017 | 0.378 |
| 2018 | 0.421 |
|  |  |
|  |  |
|  |  |

Table 4.4.8.4 Norwegian spring-spawning herring. Relative standard error of Fleet 5 used by XSAM.

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 0.206 | 0.139 | 0.157 | 0.198 | 0.243 | 0.35 | 0.776 | 0.912 | 0.443 | 0.22 |
| 1997 | 0.276 | 0.213 | 0.145 | 0.156 | 0.232 | 0.251 | 0.429 | 0.521 | 0.383 | 0.223 |
| 1998 | 0.362 | 0.28 | 0.203 | 0.149 | 0.167 | 0.242 | 0.302 | 0.426 | NA | 0.333 |
| 1999 | 0.239 | 0.373 | 0.289 | 0.221 | 0.161 | 0.188 | 0.298 | 0.393 | 0.986 | 0.38 |
| 2000 | 0.267 | 0.226 | 0.5 | 0.358 | 0.27 | 0.181 | 0.194 | 0.254 | 0.389 | 0.423 |
| 2001 | 0.175 | 0.264 | 0.262 | 0.428 | 0.415 | 0.218 | 0.193 | 0.274 | 0.498 | 0.425 |
| 2002 | 0.186 | 0.169 | 0.264 | 0.304 | 0.36 | 0.298 | 0.246 | 0.232 | 0.264 | 0.435 |
| 2003 | 0.185 | 0.168 | 0.168 | 0.261 | 0.308 | 0.449 | 0.405 | 0.248 | 0.235 | 0.242 |
| 2004 | 0.259 | 0.195 | 0.159 | 0.165 | 0.282 | 0.326 | 0.523 | 0.376 | 0.363 | 0.231 |
| 2005 | 0.143 | 0.267 | 0.251 | 0.187 | 0.194 | 0.317 | 0.358 | 0.454 | 0.392 | 0.244 |
| 2006 | 0.378 | 0.154 | 0.265 | 0.244 | 0.185 | 0.182 | 0.314 | 0.31 | 0.432 | 0.239 |
| 2007 | 0.224 | 0.19 | 0.142 | 0.272 | 0.244 | 0.184 | 0.192 | 0.317 | 0.339 | 0.225 |
| 2008 | 0.319 | 0.165 | 0.175 | 0.153 | 0.26 | 0.237 | 0.198 | 0.227 | 0.346 | 0.283 |
| 2009 | 0.254 | 0.24 | 0.16 | 0.172 | 0.17 | 0.24 | 0.264 | 0.232 | 0.265 | 0.308 |
| 2010 | 0.331 | 0.235 | 0.251 | 0.175 | 0.189 | 0.192 | 0.296 | 0.295 | 0.288 | 0.302 |
| 2011 | 0.288 | 0.26 | 0.213 | 0.23 | 0.17 | 0.201 | 0.204 | 0.3 | 0.294 | 0.284 |
| 2012 | 0.224 | 0.353 | 0.315 | 0.255 | 0.23 | 0.205 | 0.243 | 0.241 | 0.38 | 0.279 |
| 2013 | 0.304 | 0.208 | 0.348 | 0.274 | 0.243 | 0.221 | 0.2 | 0.226 | 0.248 | 0.251 |
| 2014 | 0.273 | 0.302 | 0.226 | 0.307 | 0.283 | 0.228 | 0.236 | 0.225 | 0.279 | 0.265 |
| 2015 | 0.342 | 0.27 | 0.276 | 0.213 | 0.265 | 0.277 | 0.219 | 0.231 | 0.204 | 0.245 |
| 2016 | 0.213 | 0.261 | 0.238 | 0.251 | 0.228 | 0.292 | 0.278 | 0.251 | 0.248 | 0.203 |
| 2017 | 0.283 | 0.203 | 0.275 | 0.26 | 0.296 | 0.263 | 0.372 | 0.309 | 0.288 | 0.199 |
| 2018 | 0.283 | 0.242 | 0.191 | 0.289 | 0.264 | 0.328 | 0.268 | 0.369 | 0.313 | 0.196 |

Table 4.5.1.1. Norwegian spring-spawning herring. Parameter estimates of the final XSAM model fit. The estimates from last year's assessment (from October 2017) are also shown.

| Parameter | Estimate | Std. Error | CV | Estimate 2017 | Std. Error 2017 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\log \left(N_{3,1988}\right)$ | 7.072 | 0.173 | 0.024 | 7.073 | 0.168 |
| $\log \left(N_{4,1988}\right)$ | 6.606 | 0.212 | 0.032 | 6.624 | 0.205 |
| $\log \left(N_{5,1988}\right)$ | 9.577 | 0.079 | 0.008 | 9.594 | 0.076 |
| $\log \left(N_{6,1988}\right)$ | 4.792 | 0.371 | 0.077 | 4.796 | 0.363 |
| $\log \left(N_{7,1988}\right)$ | 3.474 | 0.508 | 0.146 | 3.471 | 0.494 |
| $\log \left(N_{8,1988}\right)$ | 3.132 | 0.557 | 0.178 | 3.126 | 0.538 |
| $\log \left(N_{9,1988}\right)$ | 4.079 | 0.455 | 0.112 | 4.082 | 0.444 |
| $\log \left(N_{10,1988}\right)$ | 3.28 | 0.653 | 0.199 | 3.29 | 0.638 |
| $\log \left(N_{11,1988}\right)$ | 2.989 | 0.716 | 0.239 | 3.015 | 0.691 |
| $\log \left(N_{12,1988}\right)$ | 3.479 | 0.732 | 0.21 | 3.496 | 0.711 |
| $\log \left(q_{3}^{F 1}\right)$ | -9.544 | 0.199 | 0.021 | -9.566 | 0.212 |
| $\log \left(q_{4}^{F 1}\right)$ | -8.064 | 0.14 | 0.017 | -8.119 | 0.159 |
| $\log \left(q_{5}^{F 1}\right)$ | -7.507 | 0.126 | 0.017 | -7.551 | 0.146 |
| $\log \left(q_{6}^{F 1}\right)$ | -7.31 | 0.127 | 0.017 | -7.323 | 0.145 |
| $\log \left(q_{7}^{F 1}\right)$ | -7.134 | 0.14 | 0.02 | -7.161 | 0.158 |
| $\log \left(q_{8}^{F 1}\right)$ | -6.917 | 0.103 | 0.015 | -6.945 | 0.108 |
| $\log \left(q_{2}^{F 4}\right)$ | -14.46 | 0.189 | 0.013 | -14.418 | 0.182 |
| $\log \left(q_{3}^{F 5}\right)$ | -7.597 | 0.116 | 0.015 | -7.56 | 0.117 |
| $\log \left(q_{4}^{F 5}\right)$ | -7.127 | 0.104 | 0.015 | -7.109 | 0.105 |
| $\log \left(q_{5}^{F 5}\right)$ | -6.891 | 0.102 | 0.015 | -6.892 | 0.103 |
| $\log \left(q_{6}^{F 5}\right)$ | -6.768 | 0.106 | 0.016 | -6.752 | 0.106 |
| $\log \left(q_{7}^{F 5}\right)$ | -6.693 | 0.112 | 0.017 | -6.668 | 0.112 |
| $\log \left(q_{8}^{F 5}\right)$ | -6.509 | 0.119 | 0.018 | -6.482 | 0.119 |
| $\log \left(q_{9}^{F 5}\right)$ | -6.508 | 0.133 | 0.02 | -6.46 | 0.134 |
| $\log \left(q_{10}^{F 5}\right)$ | -6.439 | 0.15 | 0.023 | -6.405 | 0.151 |
| $\log \left(q_{11}^{F 5}\right)$ | -6.438 | 0.15 | 0.023 | -6.441 | 0.152 |
| $\log \left(\sigma_{1}^{2}\right)$ | -5 | 1.486 | 0.297 | -5 | 1.422 |
| $\log \left(\sigma_{2}^{2}\right)$ | -2.651 | 0.275 | 0.104 | -2.493 | 0.246 |
| $\log \left(\sigma_{4}^{2}\right)$ | -2.108 | 0.314 | 0.149 | -2.209 | 0.322 |
| $\log \left(\sigma_{R}^{2}\right)$ | -0.09 | 0.267 | 2.973 | -0.066 | 0.269 |
| $\boldsymbol{\operatorname { l o g }}(\mathrm{h})$ | 1.581 | 0.07 | 0.044 | 1.553 | 0.072 |
| $\mu_{R}$ | 9.361 | 0.18 | 0.019 | 9.312 | 0.186 |
| $\alpha_{Y}$ | -0.535 | 0.32 | 0.598 | -0.459 | 0.303 |
| $\beta_{Y}$ | 0.803 | 0.115 | 0.144 | 0.838 | 0.11 |
| $\alpha_{2 U}$ | -1.245 | 0.176 | 0.141 | -1.234 | 0.176 |
| $\alpha_{3 U}$ | -0.615 | 0.102 | 0.165 | -0.608 | 0.103 |
| $\alpha_{4 U}$ | -0.201 | 0.066 | 0.329 | -0.203 | 0.07 |
| $\alpha_{5 U}$ | 0.054 | 0.057 | 1.054 | 0.056 | 0.061 |
| $\alpha_{6 U}$ | 0.195 | 0.061 | 0.314 | 0.19 | 0.065 |
| $\alpha_{7 U}$ | 0.261 | 0.066 | 0.251 | 0.247 | 0.069 |
| $\alpha_{8 U}$ | 0.316 | 0.072 | 0.228 | 0.32 | 0.076 |


| Parameter | Estimate | Std. Error | CV | Estimate 2017 | Std. Error 2017 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\boldsymbol{\alpha}_{\boldsymbol{9} \boldsymbol{U}}$ | 0.373 | 0.079 | 0.211 | 0.366 | 0.081 |
| $\boldsymbol{\alpha}_{\mathbf{1 0 \boldsymbol { U }}}$ | 0.425 | 0.085 | 0.2 | 0.422 | 0.087 |
| $\boldsymbol{\beta}_{\boldsymbol{U}}$ | 0.605 | 0.055 | 0.091 | 0.61 | 0.054 |

Table 4.5.1.2 Norwegian spring-spawning herring. Point estimates of Stock in numbers (millions).

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 640 | 1178 | 739 | 14435 | 120 | 32 | 23 | 59 | 27 | 20 | 32 |
| 1989 | 1168 | 248 | 950 | 619 | 11941 | 99 | 26 | 17 | 40 | 16 | 37 |
| 1990 | 4275 | 470 | 209 | 804 | 519 | 9943 | 82 | 21 | 13 | 30 | 41 |
| 1991 | 11293 | 1732 | 399 | 177 | 677 | 433 | 8297 | 67 | 17 | 10 | 57 |
| 1992 | 18521 | 4586 | 1483 | 340 | 150 | 568 | 364 | 6918 | 55 | 14 | 56 |
| 1993 | 49735 | 7525 | 3933 | 1260 | 286 | 125 | 475 | 303 | 5720 | 45 | 57 |
| 1994 | 59395 | 20202 | 6447 | 3317 | 1029 | 232 | 102 | 385 | 243 | 4529 | 79 |
| 1995 | 15537 | 24118 | 17304 | 5428 | 2606 | 774 | 179 | 80 | 298 | 183 | 3414 |
| 1996 | 5706 | 6301 | 20605 | 14477 | 4149 | 1754 | 510 | 129 | 58 | 205 | 2227 |
| 1997 | 2086 | 2309 | 5350 | 17031 | 11085 | 2804 | 1129 | 334 | 90 | 39 | 1364 |
| 1998 | 10762 | 842 | 1915 | 4300 | 12956 | 7712 | 1750 | 661 | 206 | 54 | 759 |
| 1999 | 6439 | 4346 | 693 | 1480 | 3306 | 9448 | 5368 | 1115 | 406 | 120 | 457 |
| 2000 | 33070 | 2608 | 3621 | 541 | 1129 | 2451 | 6695 | 3599 | 697 | 240 | 302 |
| 2001 | 28868 | 13404 | 2183 | 2713 | 406 | 829 | 1750 | 4567 | 2226 | 406 | 268 |
| 2002 | 11423 | 11708 | 11367 | 1740 | 1994 | 303 | 615 | 1260 | 3165 | 1471 | 447 |
| 2003 | 6582 | 4626 | 9891 | 9175 | 1282 | 1395 | 220 | 431 | 853 | 2093 | 1282 |
| 2004 | 57638 | 2669 | 3919 | 8171 | 7204 | 945 | 1018 | 160 | 303 | 574 | 2214 |
| 2005 | 24130 | 23391 | 2268 | 3264 | 6599 | 5552 | 703 | 737 | 116 | 212 | 1736 |
| 2006 | 42853 | 9787 | 19783 | 1868 | 2605 | 5043 | 3937 | 479 | 497 | 76 | 1131 |
| 2007 | 11871 | 17381 | 8322 | 16368 | 1501 | 2035 | 3700 | 2710 | 330 | 343 | 711 |
| 2008 | 17281 | 4808 | 14743 | 6853 | 12594 | 1137 | 1488 | 2523 | 1795 | 221 | 723 |
| 2009 | 6603 | 6972 | 4067 | 12142 | 5303 | 8812 | 803 | 1022 | 1608 | 1129 | 631 |
| 2010 | 4053 | 2648 | 5832 | 3333 | 9387 | 3780 | 5726 | 536 | 633 | 953 | 1084 |
| 2011 | 15792 | 1625 | 2203 | 4781 | 2647 | 7071 | 2634 | 3568 | 335 | 387 | 1098 |
| 2012 | 4658 | 6341 | 1354 | 1801 | 3838 | 2062 | 5318 | 1791 | 2367 | 217 | 935 |
| 2013 | 7854 | 1883 | 5307 | 1113 | 1443 | 3030 | 1575 | 3909 | 1261 | 1649 | 804 |
| 2014 | 4789 | 3181 | 1585 | 4346 | 890 | 1136 | 2353 | 1176 | 2860 | 908 | 1915 |
| 2015 | 15817 | 1943 | 2705 | 1319 | 3525 | 716 | 907 | 1846 | 899 | 2156 | 2255 |
| 2016 | 8816 | 6422 | 1658 | 2272 | 1086 | 2870 | 580 | 722 | 1451 | 694 | 3525 |
| 2017 | 7135 | 3579 | 5475 | 1385 | 1853 | 866 | 2281 | 453 | 553 | 1095 | 3263 |
| 2018 | 24928 | 2891 | 3025 | 4454 | 1082 | 1377 | 624 | 1655 | 310 | 368 | 3089 |

Table 4.5.1.3 Norwegian spring-spawning herring. Point estimates of Fishing mortality.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.05 | 0.066 | 0.028 | 0.04 | 0.05 | 0.056 | 0.156 | 0.232 | 0.352 | 0.204 | 0.204 |
| 1989 | 0.011 | 0.021 | 0.016 | 0.025 | 0.033 | 0.039 | 0.075 | 0.106 | 0.148 | 0.091 | 0.091 |
| 1990 | 0.004 | 0.013 | 0.014 | 0.023 | 0.031 | 0.031 | 0.052 | 0.074 | 0.1 | 0.07 | 0.07 |
| 1991 | 0.001 | 0.005 | 0.011 | 0.018 | 0.024 | 0.025 | 0.032 | 0.043 | 0.056 | 0.043 | 0.043 |
| 1992 | 0.001 | 0.004 | 0.013 | 0.023 | 0.028 | 0.029 | 0.033 | 0.04 | 0.054 | 0.051 | 0.051 |
| 1993 | 0.001 | 0.005 | 0.02 | 0.053 | 0.06 | 0.056 | 0.062 | 0.068 | 0.083 | 0.098 | 0.098 |
| 1994 | 0.001 | 0.005 | 0.022 | 0.091 | 0.135 | 0.111 | 0.096 | 0.106 | 0.134 | 0.15 | 0.15 |
| 1995 | 0.002 | 0.007 | 0.028 | 0.119 | 0.246 | 0.268 | 0.173 | 0.171 | 0.221 | 0.329 | 0.329 |
| 1996 | 0.005 | 0.014 | 0.041 | 0.117 | 0.242 | 0.291 | 0.272 | 0.21 | 0.242 | 0.429 | 0.429 |
| 1997 | 0.007 | 0.037 | 0.068 | 0.123 | 0.213 | 0.321 | 0.385 | 0.334 | 0.358 | 0.465 | 0.465 |
| 1998 | 0.007 | 0.044 | 0.108 | 0.113 | 0.166 | 0.212 | 0.301 | 0.337 | 0.393 | 0.426 | 0.426 |
| 1999 | 0.004 | 0.033 | 0.097 | 0.121 | 0.149 | 0.194 | 0.25 | 0.32 | 0.376 | 0.497 | 0.497 |
| 2000 | 0.003 | 0.028 | 0.139 | 0.139 | 0.159 | 0.187 | 0.232 | 0.33 | 0.389 | 0.555 | 0.555 |
| 2001 | 0.003 | 0.015 | 0.077 | 0.158 | 0.14 | 0.149 | 0.179 | 0.217 | 0.264 | 0.262 | 0.262 |
| 2002 | 0.004 | 0.019 | 0.064 | 0.155 | 0.208 | 0.172 | 0.205 | 0.24 | 0.263 | 0.253 | 0.253 |
| 2003 | 0.003 | 0.016 | 0.041 | 0.092 | 0.155 | 0.165 | 0.168 | 0.203 | 0.246 | 0.272 | 0.272 |
| 2004 | 0.002 | 0.013 | 0.033 | 0.064 | 0.111 | 0.145 | 0.173 | 0.175 | 0.204 | 0.324 | 0.324 |
| 2005 | 0.002 | 0.018 | 0.044 | 0.075 | 0.119 | 0.194 | 0.234 | 0.244 | 0.268 | 0.394 | 0.394 |
| 2006 | 0.002 | 0.012 | 0.039 | 0.069 | 0.097 | 0.16 | 0.223 | 0.223 | 0.222 | 0.379 | 0.379 |
| 2007 | 0.004 | 0.015 | 0.044 | 0.112 | 0.128 | 0.163 | 0.233 | 0.262 | 0.249 | 0.227 | 0.227 |
| 2008 | 0.008 | 0.017 | 0.044 | 0.106 | 0.207 | 0.198 | 0.226 | 0.301 | 0.314 | 0.253 | 0.253 |
| 2009 | 0.014 | 0.028 | 0.049 | 0.107 | 0.189 | 0.281 | 0.254 | 0.329 | 0.373 | 0.334 | 0.334 |
| 2010 | 0.014 | 0.034 | 0.049 | 0.08 | 0.133 | 0.211 | 0.323 | 0.322 | 0.343 | 0.468 | 0.468 |
| 2011 | 0.012 | 0.032 | 0.051 | 0.07 | 0.1 | 0.135 | 0.236 | 0.26 | 0.285 | 0.313 | 0.313 |
| 2012 | 0.006 | 0.028 | 0.046 | 0.072 | 0.086 | 0.119 | 0.158 | 0.2 | 0.212 | 0.209 | 0.209 |
| 2013 | 0.004 | 0.022 | 0.05 | 0.074 | 0.089 | 0.103 | 0.143 | 0.163 | 0.179 | 0.097 | 0.097 |
| 2014 | 0.002 | 0.012 | 0.034 | 0.059 | 0.067 | 0.076 | 0.093 | 0.118 | 0.133 | 0.074 | 0.074 |
| 2015 | 0.001 | 0.009 | 0.024 | 0.044 | 0.056 | 0.062 | 0.077 | 0.091 | 0.109 | 0.074 | 0.074 |
| 2016 | 0.002 | 0.01 | 0.03 | 0.054 | 0.077 | 0.08 | 0.096 | 0.117 | 0.132 | 0.107 | 0.107 |
| 2017 | 0.003 | 0.018 | 0.057 | 0.097 | 0.147 | 0.177 | 0.171 | 0.23 | 0.258 | 0.194 | 0.194 |
| 2018 | 0.003 | 0.017 | 0.052 | 0.093 | 0.139 | 0.164 | 0.172 | 0.219 | 0.244 | 0.184 | 0.184 |

Table 4.5.1.4 Norwegian spring spawning herring. Final stock summary table. High and low represent approximate $95 \%$ confidence limits.

| Year | Recruitment <br> (Age 2) | High | Low | Stock <br> Size: <br> SSB | High | Low | Catches | Fishing Pressure: F | High | Low |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | MILLIONS |  |  | THousnd |  |  | THOUSAND | Ages 5- |  |  |
|  |  |  |  | TONNES |  |  | TONNES | 12 |  |  |
| 1988 | 640 | 338 | 942 | 2108 | 1794 | 2422 | 135 | 0.042 | 0.022 | 0.062 |
| 1989 | 1168 | 687 | 1649 | 3260 | 2774 | 3747 | 104 | 0.034 | 0.017 | 0.05 |
| 1990 | 4275 | 3179 | 5371 | 3528 | 3013 | 4043 | 86 | 0.031 | 0.016 | 0.046 |
| 1991 | 11293 | 9162 | 13423 | 3303 | 2822 | 3783 | 85 | 0.031 | 0.016 | 0.046 |
| 1992 | 18521 | 15447 | 21596 | 3331 | 2872 | 3789 | 104 | 0.038 | 0.021 | 0.056 |
| 1993 | 49735 | 43368 | 56103 | 3302 | 2890 | 3714 | 232 | 0.076 | 0.048 | 0.104 |
| 1994 | 59395 | 52269 | 66520 | 3431 | 3022 | 3841 | 479 | 0.125 | 0.089 | 0.161 |
| 1995 | 15537 | 12910 | 18163 | 3508 | 3114 | 3902 | 906 | 0.215 | 0.167 | 0.263 |
| 1996 | 5706 | 4485 | 6927 | 4096 | 3696 | 4496 | 1220 | 0.188 | 0.152 | 0.225 |
| 1997 | 2086 | 1518 | 2655 | 5355 | 4873 | 5836 | 1427 | 0.195 | 0.16 | 0.229 |
| 1998 | 10762 | 8793 | 12731 | 5908 | 5378 | 6438 | 1223 | 0.192 | 0.156 | 0.228 |
| 1999 | 6439 | 5110 | 7768 | 5770 | 5219 | 6322 | 1235 | 0.214 | 0.173 | 0.256 |
| 2000 | 33070 | 28460 | 37680 | 4799 | 4296 | 5303 | 1207 | 0.257 | 0.205 | 0.309 |
| 2001 | 28868 | 24671 | 33066 | 3986 | 3535 | 4437 | 766 | 0.203 | 0.159 | 0.248 |
| 2002 | 11423 | 9310 | 13536 | 3528 | 3109 | 3946 | 808 | 0.226 | 0.176 | 0.276 |
| 2003 | 6582 | 5193 | 7972 | 4172 | 3707 | 4637 | 790 | 0.151 | 0.118 | 0.184 |
| 2004 | 57638 | 50230 | 65046 | 5270 | 4706 | 5834 | 794 | 0.127 | 0.099 | 0.155 |
| 2005 | 24130 | 20221 | 28038 | 5401 | 4810 | 5993 | 1003 | 0.172 | 0.135 | 0.208 |
| 2006 | 42853 | 36496 | 49210 | 5365 | 4783 | 5947 | 969 | 0.175 | 0.136 | 0.215 |
| 2007 | 11871 | 9462 | 14280 | 6901 | 6176 | 7627 | 1267 | 0.153 | 0.12 | 0.186 |
| 2008 | 17281 | 13971 | 20591 | 6987 | 6215 | 7759 | 1546 | 0.2 | 0.158 | 0.242 |
| 2009 | 6603 | 5061 | 8146 | 6956 | 6128 | 7784 | 1687 | 0.207 | 0.165 | 0.249 |
| 2010 | 4053 | 2955 | 5151 | 6149 | 5338 | 6960 | 1457 | 0.217 | 0.169 | 0.264 |
| 2011 | 15792 | 12222 | 19361 | 5774 | 4938 | 6610 | 993 | 0.163 | 0.125 | 0.2 |
| 2012 | 4658 | 3283 | 6033 | 5544 | 4684 | 6404 | 826 | 0.144 | 0.109 | 0.179 |
| 2013 | 7854 | 5529 | 10178 | 5158 | 4320 | 5997 | 685 | 0.125 | 0.092 | 0.158 |
| 2014 | 4789 | 3038 | 6539 | 4924 | 4091 | 5757 | 461 | 0.087 | 0.063 | 0.11 |
| 2015 | 15817 | 10382 | 21253 | 4615 | 3811 | 5419 | 329 | 0.071 | 0.05 | 0.092 |
| 2016 | 8816 | 4504 | 13129 | 4336 | 3577 | 5095 | 383 | 0.092 | 0.065 | 0.12 |
| 2017 | 7135 | 2158 | 12112 | 4235 | 3485 | 4985 | 722 | 0.174 | 0.123 | 0.224 |
| 2018 | 24928 | 0 | 57788 | 3826 | 3065 | 4587 |  |  |  |  |
| Average | 16765 | 13046 | 20741 | 4672 | 4072 | 5271 | 798 | 0.144 | 0.110 | 0.178 |

Table 4.8.1.1 Norwegian Spring-spawning herring. Input to short-term prediction. Stock size is in millions and weight in kg.

| InPUT FOR | 2018 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Stockno. | Natural | MATURITY | Proportion of M | Proportion of F | Weight | Exploitation | Weight |
| AGE | 1-JAN. | MORTALITY | ogive | before Spawning | before Spawning | in stock | Pattern | in CATCH |
| 2 | 24928 | 0.9 | 0 | 0 | 0 | 0.054 | 0.003 | 0.133 |
| 3 | 2891 | 0.15 | 0 | 0 | 0 | 0.115 | 0.014 | 0.207 |
| 4 | 3025 | 0.15 | 0.4 | 0 | 0 | 0.149 | 0.043 | 0.256 |
| 5 | 4454 | 0.15 | 0.8 | 0 | 0 | 0.225 | 0.076 | 0.301 |
| 6 | 1082 | 0.15 | 1 | 0 | 0 | 0.226 | 0.114 | 0.328 |
| 7 | 1377 | 0.15 | 1 | 0 | 0 | 0.289 | 0.135 | 0.349 |
| 8 | 624 | 0.15 | 1 | 0 | 0 | 0.312 | 0.142 | 0.364 |
| 9 | 1655 | 0.15 | 1 | 0 | 0 | 0.343 | 0.18 | 0.374 |
| 10 | 310 | 0.15 | 1 | 0 | 0 | 0.359 | 0.201 | 0.382 |
| 11 | 368 | 0.15 | 1 | 0 | 0 | 0.361 | 0.152 | 0.384 |
| 12 | 3089 | 0.15 | 1 | 0 | 0 | 0.375 | 0.152 | 0.389 |
| INPUT FOR | 2019 AND 2020 |  |  |  |  |  |  |  |
|  | Stockno. | Natural | MATURITY | Proportion of M | Proportion of F | Weight | Exploitation | Weicht |
| AGE | 1-Jan. | MORTALITY | ogive | before spawning | before Spawning | in stock | Pattern | in CATCH |
| 2 | 11621 | 0.9 | 0 | 0 | 0 | 0.054 | 0.014 | 0.133 |
| 3 |  | 0.15 | 0 | 0 | 0 | 0.115 | 0.071 | 0.207 |
| 4 |  | 0.15 | 0.4 | 0 | 0 | 0.175 | 0.21 | 0.256 |
| 5 |  | 0.15 | 0.8 | 0 | 0 | 0.24 | 0.385 | 0.301 |
| 6 |  | 0.15 | 1 | 0 | 0 | 0.278 | 0.565 | 0.328 |
| 7 |  | 0.15 | 1 | 0 | 0 | 0.31 | 0.669 | 0.349 |
| 8 |  | 0.15 | 1 | 0 | 0 | 0.328 | 0.726 | 0.364 |
| 9 |  | 0.15 | 1 | 0 | 0 | 0.349 | $0.888$ | 0.374 |
| 10 |  | 0.15 | 1 | 0 | 0 | 0.357 | 1 | 0.382 |
| 11 |  | 0.15 | 1 | 0 | 0 | 0.358 | 0.855 | 0.384 |
| 12 |  | 0.15 | 1 | 0 | 0 | 0.374 | 0.855 | 0.389 |

Table 4.8.2.1 Norwegian spring spawning herring. Short-term prediction.

| BASIS: |  |
| :--- | :--- |
| SSB (2018): | $3.826(3.065,4.587) *$ million t |
| Landings(2018): | $546448 \mathrm{t}(\mathrm{sum}$ of national quotas)  <br> SSB(2019): $3.859(3.069,4.866)^{*}$ million t <br> Fw5-11 (2018): $0.117(0.030,0.275)^{*}$ <br> Fw5-12(2018) $0.125(0.035,0.280)^{*}$ <br> Recruitment(2018-2020): $24.928(0,57.788)^{*}, 11.621(1.009,48.205)^{*}, 11.621(1.009,48.205)^{*}$ |

The catch options:

| Rationale | $\begin{aligned} & \text { CATCHE } \\ & \text { S } \\ & (2019) \end{aligned}$ | BASIS | FW(2019) | SSB2020 | $\begin{gathered} \text { P(SSB202 } \\ 0 \\ \text { <BLIM) } \end{gathered}$ | \% SSB <br> CHANGE | \%TAC <br> CHANGE** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Zero Catch | 0 | $\mathrm{F}=0$ | 0 | $\begin{aligned} & 4.510 \\ & (3.468,6.056 \\ & )^{*} \end{aligned}$ | 0 | $\begin{aligned} & 17 \\ & (3,52)^{*} \end{aligned}$ | -100 |
| Status quo | 530319 | $\begin{aligned} & \mathrm{F}=0.1 \\ & 25 \end{aligned}$ | $\begin{aligned} & 0.125 \\ & (0.099,0.165 \\ & )^{*} \end{aligned}$ | $\begin{aligned} & 4.065 \\ & (3.050,5.552 \\ & )^{*} \end{aligned}$ | 0.001 | $5(-9,36)^{*}$ | -3 |
| MANAGENENT PLAN 19992017 | 420197 | $\begin{aligned} & \mathrm{F}=0.0 \\ & 91^{* *} \end{aligned}$ | $\begin{aligned} & \hline 0.091^{* *} \\ & (0.053,0.12)^{*} \end{aligned}$ | $\begin{aligned} & 4.157 \\ & (3.126,5.883 \\ & )^{*} \end{aligned}$ | 0 | $8(-6,44)^{*}$ | -23 |
| $F=0.085$ | 367038 | $\begin{aligned} & \mathrm{F}=0.0 \\ & 85 \end{aligned}$ | $\begin{aligned} & 0.085 \\ & (0.067,0.109 \\ & )^{*} \end{aligned}$ | $\begin{aligned} & 4.202 \\ & (3.170,5.711 \\ & )^{*} \\ & \hline \end{aligned}$ | 0 | $9(-5,42)^{*}$ | -33 |
| $\mathrm{F}=0.125^{* * *}$ | 529333 | $\begin{aligned} & \mathrm{F}=0.1 \\ & 25 \end{aligned}$ | $\begin{aligned} & 0.125 \\ & (0.099,0.161 \\ & )^{*} \end{aligned}$ | $\begin{aligned} & 4.066 \\ & (3.099,5.581 \\ & )^{*} \end{aligned}$ | 0 | $5(-9,39)^{*}$ | -3 |
| $F=0.157$ | 654642 | $\begin{aligned} & \mathrm{F}=0.1 \\ & 57 \end{aligned}$ | $\begin{aligned} & 0.157 \\ & (0.126,0.205 \\ & )^{*} \end{aligned}$ | $\begin{aligned} & 3.962 \\ & (2.950,5.387 \\ & )^{*} \end{aligned}$ | 0 | $\begin{aligned} & 2(- \\ & 12,35)^{*} \end{aligned}$ | 20 |
| $S^{\text {S }}{ }_{2020}=B_{\text {PA }}$ | 1598052 | $\begin{aligned} & \mathrm{F}=0.4 \\ & 36 \end{aligned}$ | $\begin{aligned} & \hline 0.436 \\ & (0.341,0.652 \\ & )^{*} \end{aligned}$ | $\begin{aligned} & 3.184 \\ & (2.114,4.726 \\ & )^{*} \end{aligned}$ | 0.124 | $\begin{aligned} & -18(- \\ & 35,13) \end{aligned}$ | 192 |
| $S S B_{2020}=B_{\text {LIM }}$ | 2449509 | $\begin{aligned} & \mathrm{F}=0.7 \\ & 71 \end{aligned}$ | $\begin{aligned} & 0.771 \\ & (0.593,1.360 \\ & )^{*} \end{aligned}$ | $\begin{aligned} & 2.500 \\ & (1.450,4.106 \\ & )^{*} \\ & \hline \end{aligned}$ | 0.539 | $\begin{aligned} & -35(-55,- \\ & 2)^{*} \end{aligned}$ | 348 |

[^4]
### 4.17 Figures



Figure 4.2.1.1. Total reported landings (ICES estimates) of Norwegian spring-spawning herring in 2017 by ICES rectangle. Landings below 10 tonnes per statistical rectangle are not included. The landings with information on statistical rectangle constitute $\mathbf{9 9 \%}$ of the reported landings.


Figure 4.2.1.2. Total reported landings (ICES estimates) of Norwegian spring-spawning herring in 2017 by quarter and ICES rectangle. Landings below 10 tonnes per statistical rectangle are not included. The landings with information on statistical rectangle constitute $\mathbf{9 9 \%}$ of the reported landings.


Figure 4.4.3.1. Norwegian spring spawning herring. Age disaggregated catch in numbers plotted on $a \log$ scale. Age is on x -axis. The labels indicate year classes and grey lines correspond to $\mathrm{Z}=0.3$.


Figure 4.4.4.1.Norwegian spring spawning herring. Mean weight at age by age groups $3-14$ in the years 1981-2017 in the catch (weight at age for zero catch numbers were omitted)


Figure 4.4.4.2.Norwegian spring-spawning herring. Mean weight at age in the stock 1981-2018.


Figure 4.4.5.1. Assumed (blue line) and updated (orange line) maturity-at-age for the years 2012 and 2013.


Figure 4.4.7.1.Distribution of Norwegian spring-spawning herring as measured during the IESNS survey in April-June 2018 in terms of NASC values ( $\mathrm{m}^{2} / \mathrm{nm}^{2}$ ) for every 1 nautical mile. The stratification of the survey area is shown on the map.


Figure 4.4.7.2. Norwegian acoustic survey on the NSSH spawning grounds. Distribution and acoustic density of herring recorded in 2018.


Figure 4.4.7.3. Norwegian spring spawning herring. Age disaggregated abundance indices (billions) from the acoustic survey on the spawning area in February-March (survey 1) plotted on a log scale. The labels indicate year classes and grey lines correspond to $\mathrm{Z}=0.3$. Age is on x -axis. The labels indicate year classes and grey lines correspond to $\mathrm{Z}=0.3$.


Figure 4.4.7.4. Norwegian spring spawning herring. Age disaggregated abundance indices (billions) from the acoustic survey on the feeding area in the Norwegian Sea in May (survey 5) plotted on a $\log$ scale. The labels indicate year classes and grey lines correspond to $\mathrm{Z}=0.3$.


Figure 4.5.1.1.Estimated exploitation pattern for the years 1988-2018 by the XSAM model fit. All panels shows includes the same data, but shown at different angles to improve visibility at different time periods


Figure 4.5.1.2. Norwegian spring spawning herring. Correlation between estimated parameters in the final XSAM model fit.


Figure 4.5.1.3. Norwegian spring spawning herring. Weights (inverse of variance) of data-input of the final XSAM model fit.


Figure 4.5.1.4. Norwegian spring spawning herring. Standardized residuals type 1 (left) and type 2 (right) (see text) of data-input of the final XSAM model fit.


Figure 4.5.1.5. Norwegian spring spawning herring. Observed vs. predicted values (left column) and qq-plot based on type 1 (middle) and type 2 (right) residuals (see text) based on the final XSAM model fit.


Figure 4.5.1.6. Norwegian spring spawning herring. Profiles of marginal log-likelihood $\mathrm{l}_{\mathrm{M}}$, the catch component $l_{C}$, Fleet 1 component $l_{F 1}$, Fleet 4 component $l_{F 4}$, Fleet 5 component $l_{F 5}$, point estimate of SSB and average $F$ (ages 5-12+) in 2017 over the common scaling factor for variance in data $h$ for the final XSAM fit. The red dots indicate the value of the respective scaling factors for which the log-likelihood is maximized.


Figure 4.5.1.7. Norwegian spring spawning herring. Retrospective XSAM model fits of SSB and weighted average of fishing mortality ages 5-11 for the years 2012-2017.


Figure 4.5.1.8. Norwegian spring spawning herring. Point estimates of Spawning-stock biomass by years 1988-2018 from model (black lines) and by survey indices from Fleet 1 (red) and Fleet 5 (green). Dotted lines are approximate $95 \%$ confidence interval.


Figure 4.5.1.9. Total reported landings 1988-2017, estimated recruitment, weighted average of fishing mortality (ages 5-12) and spawning-stock biomass for the years 1988-2018 based on the final XSAM model fit. The broken lines are approximate $95 \%$ confidence limits.


Figure 4.5.2.1.1. Norwegian spring spawning herring. Q-Q plot from the eight different surveys used in tuning in TASACS. First row starts with survey 1 and the last one in row four is larval survey.


Figure 4.5.2.1.2. Norwegian spring-spawning herring. Residual sum of squares in the surveys separately from TASACS. First row starts with survey 1 and the last one in row four is larval survey.


Figure 4.5.2.1.3. Comparison of SSB time-series from the final assessment from XSAM and exploratory runs from TASACS (following the 2008 benchmark procedure) and TISVPA. $95 \%$ confidence intervals from the XSAM final assessment are shown.


Figure 4.8.1.1. Estimated selection pattern by XSAM; thin grey lines shows annual estimates 19882017, the median value is indicated by the thick grey line, while selected years (estimates for 20142017 and predictions for 2018-2019) are shown in colours as indicated in the legend.


Figure 4.9.1. Norwegian spring spawning herring. Comparisons of spawning stock; weighted fishing mortality $F(5-14)$ and $F(5-11)$; and recruitment at age 0 and age 2 with previous assessments. In 2016 the proportion mature in the years 2006-2011 was changed; recruitment age changed from 0 to 2 and fishing mortality is calculated over ages 5 to 11. In 2018 (WKNSSHREF) the age range for the fishing mortality changed to ages 5-12.

## 5 Horse Mackerel in the Northeast Atlantic

### 5.1 Fisheries in 2017

The total international catches of horse mackerel in the North East Atlantic are shown in Table 5.1.1 and Figure 5.4.1. The southern horse mackerel stock is currently assessed by ICES WGHANSA. The total catch from all areas in 2017 for the Western and North Sea stock was 97,540 tons which is 16611 tons less than in 2016 (and $12 \%$ lower than in 2015). France and the Netherlands have a directed trawl fishery and Norway and France a directed purse-seine fishery for horse mackerel. Spain has directed and mixed trawl and purse-seine fisheries. In earlier years most of the catches were used for meal and oil while in later years most of the catches have been used for human consumption.

The quarterly catches of North Sea and western horse mackerel by Division and Subdivision in 2017 are given in Table 5.1.2 and the distributions of the fisheries are given in Figure 5.1.1.a-d. The maps are based on data provided by Belgium, Faroe Islands, France, Germany, Ireland, Netherlands, Norway, Sweden, Spain and UK (Engl. And Wales) representing $99 \%$ of the total catches. The distribution of the fishery is similar to the later years.

The Dutch, Danish, Irish and German fleets operated mainly in the North and West of Ireland and the Western waters off Scotland. The French fleet were in the Bay of Biscay and West Scotland whereas the Norwegian fleet fished in the North-eastern part of the North Sea. The Spanish fleet operated mainly in waters of Cantabrian Sea and Bay of Biscay.
First quarter: The fishing season with most of the catches 39,251 tons ( $47 \%$ of the total catches). The fishery was mainly carried out west of Scotland and West and North of Ireland and along the Spanish coast (Figure 5.1.1.a).
Second quarter: 7,377 tons. As usual, catches were significantly lower than in the first quarter as the second quarter is the main spawning period. Most of the catches were taken West of Ireland and along the Spanish coast. (Figure 5.1.1.b)

Third quarter: 12,921 tons. Most of the catches were taken in Spanish waters and at the Norwegian coast. Also some smaller catches were reported in the Southern part of the North Sea (Figure 5.1.1.c).

Fourth quarter: Catches were 23,381 tons. The catches were distributed in four main areas (Figure 5.1.1.d):

- Spanish waters,
- Northern Irish waters and West of Scotland
- Norwegian coast
- East part of Channel


### 5.2 Stock Units

For many years the Working Group has considered the horse mackerel in the Northeast Atlantic as separated into three stocks: the North Sea, the Southern and the Western stocks (ICES 1990, ICES 1991). For further information see Stock Annex Western Horse Mackerel and to the WD document on horse mackerel stock structure (WD Brunel et al., 2016). The boundaries for the different stocks are given in Figure 5.2.1.

To improve on the understanding of the stock structure, horse mackerel samples for genetic analysis have been collected in the central and Northern North Sea, Channel, West of Ireland, the Bay of Biscay, Cantabrian Sea and in the waters around Morocco and Mauritania (as out-group). Samples have been collected mostly during spawning time in the years 2015 to 2017. It is foreseen that the genetic analysis will be carried out in 2018 leading to potential results before the next WGWIDE in 2019.

### 5.3 WG Catch Estimates

In 2017, a review of catch statistics for North Sea and Western horse mackerel stocks was carried out. The results of this report have been reported in previous Working Groups reports. (Costas, 2017)
As a result of this review catches and catch-at-ages of reported historical data of both North Sea and Western stocks of horse mackerel were updated. Catch statistics were reviewed since 1990 onward for Western stock and since 2000 onward for North Sea stock. Main mismatches between the catch statistics in working group reports and these reviewed data were originated by several reasons such as late availability of some data for the report or the availability of only official catch figures.

### 5.4 Allocation of Catches to Stocks

The distribution areas for the three stocks are given in the Stock Annex for the Western Horse Mackerel. The catches in 2017 were allocated to the three stocks as follows:

Western stock: 3 and 4 quarter: Divisions 3.a and 4.a. 1-4 quarter: 2.a, 5.b, 6.a, 7.a-c, ek and 8.a-e.

North Sea stock: 1 and 2 quarter: Divisions 3.a and 4.a 1-4 quarter: Divisions 4.b, 4.c and 7.d.

Southern stock: Division 9.a. All catches from these areas were allocated to the southern stock. This stock is now dealt with by another working group (ICES WGHANSA).
The catches by stock are given in Table 5.4.1 and Figure 5.4.1. The catches by ICES subArea and division for the Western and North Sea stocks for period 1992-2017 are shown in Figures 5.4.2-3. The catches by stock and countries for the period 1997-2017 are given in Table 5.4.2-5.4.3.

### 5.5 Estimates of discards

Over the years only Netherlands had provided data on discards and in some few years also Germany and Spain. For 2017 almost all of countries provided such data. The provided discard rate is less than $5.3 \%$ in weight for the combined Horse mackerel stocks. The discard rate for the North Sea stock is estimated to be $8.3 \%$ and for the Western stock $4.7 \%$ in 2017.

### 5.6 Trachurus Species Mixing

Three species of genus Trachurus: T. trachurus, T. mediterraneus and T. picturatus are found together and are commercially exploited in NE Atlantic waters. Following the Working Group recommendation (ICES 2002/ACFM: 06) special care was taken to ensure that catch and length distributions and numbers-at-age of T. trachurus supplied to the Working Group did not include T. mediterraneus and/or T. picturatus.
T. mediterraneus fishery takes mainly place in the eastern part of ICES Division 8.c. There is not a clear trend in T. mediterraneus catches in this area but in the last year's
show a low level (Table 5.6.1). Information of T. picturatus fishery is available in the WGHANSA Report (Working Group on Horse Mackerel, Anchovy and Sardine).

Taking into account that the assessment is only made for T. trachurus, the Working Group recommends that the TACs and any other management regulations which might be established in the future should be related only to T. trachurus and not to Trachurus spp. More information is needed about the Trachurus spp. before the fishery and the stock can be evaluated.

### 5.7 Length Distribution by Fleet and by Country:

Ireland, Germany, Netherlands, Norway, France, Scotland and Spain provided length distributions for their catches in 2017. The length distributions are covering app. 97\% of the total landings of the Western and North Sea horse mackerel catches and are shown in Table 5.7.1.

### 5.8 Comparing trends between areas and stocks

Horse mackerel (Trachurus trachurus) in the northeast Atlantic is assumed to be separate into three stocks:

- North Sea (4a part of the year, $4 \mathrm{~b}, 4 \mathrm{c}$ and 7 d )
- Western (4a part of the year, 5b, 6a, 7a-c,e-k, 8a-d)
- Southern (9a)

Catches in biomass between 2000 and 2017 are shown in figure 5.8.1 indicated an overall decline in the catches of horse mackerel, but with a relative increase in southern horse mackerel in the recent years.
The catch in numbers by age groups $0-3$ (juveniles), 4-10 (adults), 11-15 (seniors) are shown in figure 5.8.2. The values are indicating an increase in the catches of juveniles in the Western and North Sea stocks in recent years. This could be an indication of a stronger recruitment of horse mackerel which has been reported by surveys and fishermen. However, it is also an alarming signal if a larger proportion of the catch consists of juveniles.
The relative catch in numbers by stock, age, year and cohort are shown in figure 5.8.3. This type of display allows the cohorts to be followed through the ages and years. The strong 2001 year class clearly stands out alone in the Western stock whereas in the North Sea stock the same year class and the surrounding year classes seem to be relatively strong. Year classes in the Southern area are less clearly identified which could be due to the fishery concentrating on the younger year classes.

The relative catch in numbers by stock/area, age, year and cohort are shown in figure 5.8.4. The strong 2001 year class is most noticeable in area 6 and 7 and for the younger ages in area 8 . The 2001 year class is not very apparent in the western stock in 4 a . For the North Sea stock, the cohort signal is only apparent in area 7 and not in area 4.

The catch in number by area and age from sampled catches is shown in figure 5.8.5. There appears to be a very limited sampling for horse mackerel in area 8a in the recent year even though there are sizeable catches in that area, predominantly believed to be of younger ages. Also in area 7.h there has been no sampling in 2016. An important signal to be derived from these plots is that there appears to be an increase in the catches of juveniles in the most recent years, mostly in area 7.d and to a lesser extend
also in area 7e. Measures to protect the incoming year classes of these species should be considered.

### 5.9 Quality and Adequacy of fishery and sampling data

Table 5.9.1 shows a summary of the overall sampling intensity on horse mackerel catches in recent years in all areas 1992-2017 and in the Western and North Sea stock areas for the following years. Since 2009 the Southern horse mackerel is dealt with by ICES WGHANSA.

Countries that usually carried out sampling were Ireland, the Netherlands, Germany, Norway and Spain and they covered $42-100 \%$ of their respective catches. In 2017 Denmark, France, Germany, Ireland, the Netherlands, England, Scotland and Spain provided samples and length distributions and Germany, Ireland, the Netherlands, and Spain provided also age distributions. However, the lack of age distribution data for relatively large portions of the horse mackerel catches continues to have a serious effect on the accuracy and reliability of the assessment and the Working Group remain especially concerned about the low number of fish which are aged.

Table 5.9.2 shows the sampling intensity for the Western stock in 2017, table 5.9.3 shows the sampling intensity for the North Sea stock in 2017

An analysis on the sampling intensity was carried out for the was made analyzing sampling intensity in period 2000-2017 for both the North Sea and the Western stock in last WIDE meeting (Costas, 2017b). Sampling intensity in fisheries can be defined as the ratio of sampled catch to the total catch. The precision and accuracy of sampled catch are considerable importance to obtain a reliable estimate of the commercial catch. Sampled catch is used to extrapolate to total catch in order to obtain a catch-at-age (length) and weight at age which are often used as inputs for the stock assessment models. In addition, in case of horse mackerel the impact of temporal (quarter) and spatial (area by ICES division) factors have to be taken in account in order to obtain a reliable estimate of the commercial catches.

Figure 5.9.1 shows the proportion of sampled catches by division for the North Sea stock. In general all ICES divisions show low levels of sampling especially in the last years. The sampling intensity in relation to the length composition of catch was $62 \%$ but in relation to age composition around $39 \%$ in 2017 (Figure 5.9.2). In addition, divisions that are usually not sampled can be affect the precision and accuracy of total catch-at-age and weight at age. Figures 5.9 .3 show ratio of numbers of individuals and otoliths taken to characterize the length and age composition by 1000 t of the commercial horse mackerel catches from the North Sea. These estimates can be biased, however, since samples are usually less than the recommended 100 fish/sample. (Table 5.9.1)

The proportion of the sampled catches by region for the Western stock are showed in figure 5.9.5. Most of the regions present an adequate level of sampling although the Biscay and Channel regions show low levels of sampling in the last years. The general index of sampling intensity is around $63 \%$, although divisions (regions) that are not sampled can affect the precision and accuracy of total catch-at-age and weight at age (Figure 9.5.6). Figures 5.9.7-8 show the ratio of numbers of individuals and otoliths taken to characterize the length and age composition by 1000 t of the commercial catches. These estimates can be biased, however, since samples are usually less than the recommended 100 fish/sample. (Table 5.9.1). It has been a significant increase in
number of measured individuals per 1000 t in 2016 and 2017 produced by large increase of number of sampled individuals in division 8.b.

Length distributions were supplied by a number of countries. However, as some countries only deliver catch-at-age distributions and others only length distributions of the catch, the obtained catch-at-age and length distributions are not reflecting the total catch especially in case of North Sea horse mackerel. Furthermore, some of the length distributions are only taken from discards of non-horse mackerel targeting fleets omitting the horse mackerel targeting fleet. This lack of coverage might also have a serious effect on the accuracy and reliability of the assessment and is a matter of concern for the Working Group.

### 5.10 References

Brunel, T., 2017. Revision of the Maturity Ogive for the Western Spawning Component of NEA Mackerel. Working document to WKWIDE, 6pp.

Costas, G. 2017. Review of Horse Mackerel catch data . North Sea and Western Stocks. WD to WGWIDE 2017. 11 pp.

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### 5.11 Tables

Table 5.1.1 HORSE MACKEREL general. Catches (t) by Sub-area. Data as submitted by Working Group members. Data of limited discard information are only available for some years.

| Subarea | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 2 | - | + | - | 412 | 23 | 79 | 214 |
| 4+3.a | 1,412 | 2,151 | 7,245 | 2,788 | 4,420 | 25,987 | 24,238 | 20,746 |
| 6 | 7,791 | 8,724 | 11,134 | 6,283 | 24,881 | 31,716 | 33,025 | 20,455 |
| 7 | 43,525 | 45,697 | 34,749 | 33,478 | 40,526 | 42,952 | 39,034 | 77,628 |
| 8 | 47,155 | 37,495 | 40,073 | 22,683 | 28,223 | 25,629 | 27,740 | 43,405 |
| 9 | 37,619 | 36,903 | 35,873 | 39,726 | 48,733 | 23,178 | 20,237 | 31,159 |
| Total | 137,504 | 130,970 | 129,074 | 104,958 | 147,195 | 149,485 | 144,353 | 193,607 |
| Subarea | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| 2 | 3,311 | 6,818 | 4,809 | 11,414 | 3200 | 13457 | 0 | 759 |
| 4+3.a | 20,895 | 62,892 | 112,047 | 145,062 | 71,195 | 120,054 | 145,965 | 111,899 |
| 6 | 35,157 | 45,842 | 34,870 | 20,904 | 29,726 | 39,061 | 65,397 | 69,616 |
| 7 | 100,734 | 90,253 | 138,890 | 192,196 | 150,575 | 183,458 | 202,083 | 196,192 |
| 8 | 37,703 | 34,177 | 38,686 | 46,302 | 42,840 | 54,172 | 44,726 | 35,501 |
| 9 | 24,540 | 29,763 | 29,231 | 24,023 | 34,992 | 27,858 | 31,521 | 28,442 |
| Disc |  |  |  |  | 5,440 | 2,220 | 9,530 | 4,565 |
| Total | 222,340 | 269,745 | 358,533 | 439,901 | 337,968 | 440,280 | 499,222 | 446,974 |
| Subarea | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 2 | 13151 | 3366 | 2601 | 2544 | 2557 | 919 | 310 | 1324 |
| $4+3 . a$ | 100,916 | 25,998 | 79,761 | 34,917 | 58,745 | 31,435 | 18,513 | 52,337 |
| 6 | 83,568 | 81,311 | 40,145 | 35,073 | 40,381 | 20,735 | 24,839 | 14,843 |
| 7 | 328,995 | 263,465 | 326,469 | 300,723 | 186,622 | 140,190 | 138,428 | 98,677 |
| 8 | 28,707 | 48,360 | 40,806 | 38,571 | 48,350 | 54,197 | 75,067 | 55,897 |
| 9 | 25,147 | 20,400 | 29,491 | 41,574 | 27,733 | 26,160 | 24,912 | 23,665 |
| Disc | 2,076 | 17,082 | 168 | 996 | 0 | 385 | 254 | 307 |
| Total | 582,560 | 459,982 | 519,441 | 454,398 | 364,388 | 274,022 | 282,323 | 247,049 |
| Subarea | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 2 | 36 | 42 | 176 | 27 | 366.34 | 572 | 1847 | 1667 |
| $4+3 . a$ | 34,095 | 30,736 | 40,594 | 37,583 | 16,226 | 15,628 | 78,064 | 13,600 |
| 6 | 23,772 | 22,177 | 22,053 | 15,722 | 25,949 | 25,867 | 17,775 | 23,199 |
| 7 | 123,428 | 115,739 | 106,671 | 101,183 | 93,013 | 102,755 | 96,915 | 148,701 |
| 8 | 41,711 | 24,126 | 41,491 | 34,121 | 28,396 | 33,756 | 33,580 | 39,659 |
| 9 | 19,570 | 23,581 | 23,111 | 24,557 | 23,423 | 23,596 | 26,496 | 27,217 |
| Disc | 842 | 2,356 | 1,864 | 1,431 | 509 | 474 | 1,483 | 434 |
| Total | 243,455 | 218,758 | 235,961 | 214,624 | 187,882 | 202,649 | 256,161 | 254,478 |
| Subarea | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 20171 |  |
| 2 | 647.588 | 66.02912 | 30 | 424.291 | 10 | 45.276 | 5 |  |
| $4+3 . a$ | 25,158 | 5,234 | 8,183 | 17,270 | 10,560 | 11,565 | 12,609 |  |
| 6 | 39,496 | 44,971 | 43,266 | 32,444 | 24,153 | 32,186 | 28,170 |  |
| 7 | 120,340 | 120,476 | 100,859 | 66,853 | 49,644 | 46,901 | 33,297 |  |


| $\mathbf{8}$ | 35,245 | 17,209 | 26,983 | 30,844 | 19,822 | 17,511 | 18,307 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{9}^{3}$ | 22,575 | 25,316 | 29,382 | 29,205 | 33,179 | 41,081 | 37,080 |
| Disc | 430 | 3,279 | 4,582 | 1,904 | 6,232 | 5,944 | $5,488^{2}$ |
| Total | 243,892 | 216,552 | 213,285 | 178,945 | 143,600 | 155,232 | 134,956 |

${ }^{1}$ Preliminary. ${ }^{2}$ includes BMS of $\mathbf{1 1}$ tonnes
${ }^{3}$ Southern Horse Mackerel (ICES Division 9) is assessed by ICES WGHANSA since 2011

Table 5.1.2 HORSE MACKEREL Western and North Sea Stock combined. Quarterly catches (1000 t) by Division and Subdivision in 2017.

| Division | 1 Q | 2Q | 3Q | 4Q | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $2 . a+5 . b$ | 3 | 0 | 0 | 2 | 5 |
| 3 | + | 0 | 703 | 9 | 712 |
| 4.a | 29 | 28 | 7275 | 2129 | 9461 |
| 4.bc | 68 | 274 | 116 | 2119 | 2577 |
| 7.d | 2851 | 528 | 848 | 7719 | 11946 |
| 6.a,b | 17914 | 222 | 20 | 10061 | 28355* |
| 7.a-c,e-k | 17909 | 2162 | 632 | 2638 | 23340 |
| 8.a-e | 3425 | 4954 | 4292 | 8543 | 21213 |
| Sum | 42199 | 8167 | 13886 | 33219 | 97540 |

+ less than $50 t$, * for the total $69 t$ were added which were only declared as yearly catch

Table 5.4.1 HORSE MACKEREL general. Landings and discards (t) by year and Division, for the North Sea, Western, and Southern horse mackerel stocks. (Data submitted by Working Group members.)

| Year | $3 . \mathrm{A}$ | 4.A | 4.B,C | 7.D | DIsC | NS Sтоск | 2.A 5.B | 3.A | 4.A | 6.A,B | $\begin{gathered} \text { 7.A-C, E- } \\ K \end{gathered}$ | 8.A-E | Disc | Western Stock | W + NS <br> Stоск | $\begin{aligned} & \text { SOUTHERN } \\ & \text { STOCK }(9 . A)^{x} \end{aligned}$ | ALL Stocks |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 2,788* |  | - | 1,247 |  | 4,035 | - |  | - | 6,283 | 32,231 | 3,073 | - | 61,197 | 65,232 | 39,726 | 104,958 |
| 1983 | 4,420* |  | - | 3,600 |  | 8,020 | 412 |  | - | 24,881 | 36,926 | 28,223 | - | 90,442 | 98,462 | 48,733 | 147,195 |
| 1984 | 25,893* |  | - | 3,585 |  | 29,478 | 23 |  | 94 | 31,716 | 38,782 | 25,629 | 500 | 96,744 | 126,222 | 23,178 | 149,400 |
| 1985 | - |  | 22,897 | 2,715 |  | 26,750 | 79 |  | 203 | 33,025 | 35,296 | 27,740 | 7,500 | 103,843 | 129,455 | 20,237 | 150,830 |
| 1986 | - |  | 19,496 | 4,756 |  | 24,648 | 214 |  | 776 | 20,343 | 72,761 | 43,405 | 8,500 | 145,999 | 170,251 | 31,159 | 201,806 |
| 1987 | 1,138 |  | 9,477 | 1,721 |  | 11,634 | 3,311 |  | 11,185 | 35,197 | 99,942 | 37,703 | - | 187,338 | 199,674 | 24,540 | 223,512 |
| 1988 | 396 |  | 18,290 | 3,120 |  | 23,671 | 6,818 |  | 42,174 | 45,842 | 81,978 | 34,177 | 3,740 | 214,729 | 236,535 | 29,763 | 268,163 |
| 1989 | 436 |  | 25,830 | 6,522 |  | 33,265 | 4,809 |  | 85304** | 34,870 | 131,218 | 38,686 | 1,150 | 296,037 | 328,825 | 29,231 | 358,533 |
| 1990 | 2,261 |  | 17,437 | 1,325 |  | 18,762 | 11,414 | 14,878 | 112753** | 20,794 | 182,580 | 46,302 | 9,930 | 398,645 | 419,668 | 24,023 | 441,430 |
| 1991 | 913 | 0 | 11,400 | 600 | 0 | 12,913 | 3,200 | 2,725 | 56,157 | 29,726 | 149,975 | 42,840 | 5,440 | 290,063 | 302,976 | 34,992 | 337,968 |
| 1992 | 0 | 0 | 13,955 | 688 | 400 | 15,043 | 13,457 | 2,374 | 103,725 | 39,061 | 182,770 | 54,172 | 1,820 | 397,379 | 412,422 | 27,858 | 440,280 |
| 1993 | 0 | 0 | 3,895 | 8,792 | 930 | 13,617 | 0 | 850 | 141,220 | 65,397 | 193,291 | 44,726 | 8,600 | 454,084 | 467,701 | 31,521 | 499,222 |
| 1994 | 0 | 0 | 2,496 | 2,503 | 630 | 5,629 | 759 | 2,492 | 106,911 | 69,616 | 193,689 | 35,501 | 3,935 | 412,903 | 418,532 | 28,442 | 446,974 |
| 1995 | 112 | 0 | 7,948 | 8,666 | 30 | 16,756 | 13,151 | 128 | 92,728 | 83,568 | 320,329 | 28,707 | 2,046 | 540,657 | 557,413 | 25,147 | 582,560 |
| 1996 | 1,657 | 0 | 7,558 | 9,416 | 212 | 18,843 | 3,366 | 0 | 16,783 | 81,311 | 254,049 | 48,360 | 16,870 | 420,739 | 439,582 | 20,400 | 459,982 |
| 1997 | 0 | 0 | 14,078 | 5,452 | 10 | 19,540 | 2,601 | 2,037 | 63,646 | 40,145 | 321,017 | 40,806 | 158 | 470,410 | 489,950 | 29,491 | 519,441 |
| 1998 | 3,693 | 0 | 10,530 | 16,194 | 83 | 30,500 | 2,544 | 3,693 | 17,001 | 35,073 | 284,529 | 38,571 | 913 | 382,324 | 412,824 | 41,574 | 454,398 |
| 1999 | 0 | 0 | 9,335 | 27,889 | 0 | 37,224 | 2,557 | 2,095 | 47,315 | 40,381 | 158,733 | 48,350 | 0 | 299,431 | 336,655 | 27,733 | 364,388 |
| 2000 | 0 | 176 | 25,931 | 19,019 | 4 | 45,130 | 919 | 1,014 | 4,314 | 20,735 | 121,171 | 54,197 | 382 | 202,732 | 247,862 | 26,160 | 274,022 |
| 2001 | 43 | 212 | 6,686 | 21,390 | 0 | 28,331 | 310 | 134 | 11,438 | 24,839 | 117,038 | 75,067 | 254 | 229,081 | 257,411 | 24,912 | 282,323 |
| 2002 | 0 | 639 | 15,303 | 11,323 | 0 | 27,264 | 1,324 | 174 | 36,221 | 14,843 | 87,354 | 55,897 | 307 | 196,120 | 223,384 | 23,665 | 247,049 |
| 2003 | 49 | 622 | 10,309 | 21,049 | 0 | 32,028 | 36 | 1,843 | 21,272 | 23,772 | 102,379 | 41,711 | 842 | 191,856 | 223,885 | 19,570 | 243,455 |
| 2004 | 303 | 133 | 18,544 | 16,455 | 0 | 35,435 | 42 | 48 | 11,708 | 22,177 | 99,284 | 24,126 | 2,356 | 159,742 | 195,177 | 23,581 | 218,758 |
| 2005 | 0 | 1,331 | 13,995 | 15,460 | 62 | 30,848 | 176 | 284 | 24,983 | 22,053 | 91,211 | 41,491 | 1,802 | 182,001 | 212,850 | 23,111 | 235,961 |
| 2006 | 185 | 2,192 | 7,996 | 23,789 | 78 | 34,240 | 27 | 58 | 27,152 | 15,722 | 77,394 | 34,121 | 1,353 | 155,827 | 190,067 | 24,557 | 214,624 |
| 2007 | 11 | 2,051 | 9,114 | 29,789 | 139 | 41,103 | 366 | 110 | 4,940 | 25,949 | 63,224 | 28,396 | 370 | 123,356 | 164,459 | 23,423 | 187,882 |
| 2008 | 27 | 910 | 2,582 | 32,185 | 0 | 35,704 | 572 | 3 | 12,107 | 25,867 | 70,570 | 33,756 | 474 | 143,349 | 179,053 | 23,596 | 202,649 |
| 2009 | 21 | 314 | 18,975 | 25,537 | 1,036 | 45,883 | 1,847 | 17 | 58,738 | 17,775 | 71,378 | 33,580 | 447 | 183,782 | 229,665 | 26,496 | 256,161 |
| 2010 | 0 | 100 | 1,969 | 22,077 | 2 | 24,149 | 1,667 | 88 | 11,442 | 23,199 | 126,624 | 39,659 | 432 | 203,112 | 227,261 | 27,217 | 254,478 |
| 2011 | 0 | 0 | 10,435 | 17,184 | 0 | 27,619 | 648 | 0 | 14,723 | 39,496 | 103,156 | 35,245 | 430 | 193,698 | 221,317 | 22,575 | 243,892 |
| 2012 | 0 | 355 | 1,559 | 19,464 | 0 | 21,378 | 66 | 9 | 3,311 | 44,971 | 101,012 | 17,209 | 3,279 | 169,858 | 191,236 | 25,316 | 216,552 |
| 2013 | 0 | 17 | 1,453 | 17,175 | 0 | 18,645 | 30 | 10 | 6,702 | 43,266 | 83,684 | 26,983 | 4,582 | 165,258 | 183,903 | 29,382 | 213,285 |
| 2014 | 1 | 2 | 2,597 | 10,772 | 7 | 3,380 | 424 | 4,096 | 10,573 | 32,444 | 56,081 | 30,844 | 1,896 | 136,360 | 149,740 | 29,205 | 78,94 |


| Year | 3.A | 4.A | 4.в, С | 7.D | DIsC | $\begin{aligned} & \text { NS } \\ & \text { Stоск } \end{aligned}$ | 2.A 5.B | 3.4 | 4.A | 6.A,B | $\begin{gathered} \text { 7.A-C, E- } \\ k \end{gathered}$ | 8.A-E | DIsc | Western Stock | $\begin{aligned} & \text { W + NS } \\ & \text { Stock } \end{aligned}$ | Southern <br> Stock(9.A) ${ }^{\mathrm{x}}$ | ALL sтоскs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 3 | 644 | 770 | 8,581 | 2,004 | 12,002 | 10 | 65 | 9,078 | 24,153 | 41,063 | 19,822 | 4,228 | 98,419 | 110,421 | 33,179 | 143,600 |
| 2016 | 2 | 1,628 | 975 | 11,209 | 1,527 | 15,341 | 45 | 0 | 8,960 | 32,186 | 35,692 | 17,511 | 4,417 | 98,811 | 114,151 | 41,081 | 155,232 |
| 2017 | 0 | 22 | 2,557 | 10,787 | 1,213 | 145,79 | 5 | 697 | 9,332 | 28,170 | 22,510 | 18,307 | 3,939 | 82,961 | 97,540 | 37,088 | 134,956 |

*Divisions 3.a and 4.b,c combined• "Norwegian catches in 4.b included in Western horse mackerel. $\quad$ x Southern Horse Mackerel is assessed by ICES WGHANSA since 2011

Table 5.4.2 National catches of the Western Horse mackerel stock.


| Country | 2014 | 2015 | 2016 | 20171 |
| :--- | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  |
| Denmark | 5,945 | 4,556 | 321 | 4,541 |
| Faroe Islands | 68 | - | - | 180 |
| France | 3,428 | 3,247 | 2,797 | 3,923 |
| Germany, Fed.Rep. | 17,161 | 9,417 | 11,414 | 7,172 |
| Ireland | 32,667 | 21,654 | 27,605 | 23,560 |
| Lithuania | - | - | 2,596 | - |
| Netherlands | 25,053 | 24,958 | 23,792 | 14,269 |
| Norway | 14,353 | 8,897 | 9,438 | 9,885 |
| Spain | 19,442 | 13,071 | 14,235 | 14,901 |
| Sweden | 0 | 10 | - | 41 |
| UK (Engl. + Wales) | 4,832 | 2,063 | 842 | 549 |
| UK (N. Ireland) | 1,579 | 1,204 | - |  |
| UK (Scotland) | 1,389 | 738 | 970 | - |
| Unallocated | 8,545 | 4,377 | 1,010 | 3,994 |
| Discard | 1,896 | 4,228 | 4,417 | 3,928 |
| Total | 136,360 | 98,419 | 98,810 | 82,950 |

${ }^{1}$ Preliminary

Table 5.4.3. National catches of the North Sea Horse mackerel stock.

| Country | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 22003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | - | 19 | 21 |  |  | 30 | 5 | 4 | 4 | - |
| Denmark | 180 | 1,481 | 3,377 | 4,403 | 885 | 2,315 | 5 3,301 | 8,690 | 3,987 | 8,353 |
| Faroe Islands | - | - | 135 | - | - | 28 | 804 | 21 | - | - |
| France | 3,246 | 2,399 | - | - |  | 1,246 | 6 2,326 | 231 | 5,236 | 1,205 |
| Germany, Fed.Rep. | 7,847 | 5,844 | 5,920 | 3,728 | 974 | 6,532 | 2,936 | 5,194 | 2,725 | 11,034 |
| Ireland | - | 2,861 | 27 | 201 | 338 | 61 | - | 1 | 753 | 10,863 |
| Lithuania | - | $\begin{gathered} 10,71 \\ 1 \end{gathered}$ | - | - | - | - | - | - | - | 26,779 |
| Netherlands | 36,855 | - | 8,117 | 8,697 | $\begin{gathered} 13,86 \\ 7 \end{gathered}$ | $\begin{gathered} 12,20 \\ 9 \end{gathered}$ | $24,11$ $9$ | 26,303 | 27,730 | 6,829 |
| Norway | - | - | 238 | 105 | 36 | 525 | 144 | 22 | 204 | 37,130 |
| Sweden | - | 3,401 | 5 | 40 | 46 | 16 | 72 | 98 | 4 | 27,114 |
| UK (Engl. + Wales) | 269 | 907 | 11 | 1,585 | 3,425 | 2,322 | 2 1,966 | 5,633 | 3,859 | - |
| UK (Scotland) | 29 | - | - | 421 | - | 2 | 1 | 2 | - | 13,878 |
| Unallocated | $28,896$ | 2,794 | $\begin{gathered} 19,37 \\ 3 \end{gathered}$ | $\begin{gathered} 25,94 \\ 4 \end{gathered}$ | 8,805 | 1,981 | -3,645 | $13,064$ | $13,719$ | - |
| Discard | 10 | 83 | - | 4 | - |  | - | - | 62 | 3,583 |
| Total | 19,540 | $\begin{gathered} 30,50 \\ 0 \end{gathered}$ | $\begin{gathered} 37,22 \\ 4 \end{gathered}$ | $\begin{gathered} 45,12 \\ 8 \end{gathered}$ | $\begin{gathered} 28,37 \\ 6 \end{gathered}$ | $\begin{gathered} 27,26 \\ 7 \end{gathered}$ | $\begin{array}{cc} 6 & 32,02 \\ 9 \end{array}$ | 33,135 | 30,845 | $\begin{gathered} 155,09 \\ 4 \end{gathered}$ |
| Country | 2006 | 2007 | 2008 | 2009 | 2010 |  | 2011 | 2012 | 2013 | 2014 |
| Belgium |  |  |  | 4 | 16 |  |  | 46 | 51.077 | 74 |
| Denmark | 1,283 | 252 | 57 | 72 | 15 |  | 142 | 1514 | 1,020 | 552 |
| Faroe Islands | - | - | - | - | - |  | - | 0 |  |  |
| France | 4,380 | 5,349 | 2,247 | - | 813 |  | 273 | 1,047 | 1,010 | 1,742 |
| Germany, Fed.Rep. | 1,125 | 65 | 1,081 | 1,539 | 3,794 |  | 3,461 | 5,356 | 2,941 | 1,619 |
| Ireland | 2,077 |  | 887 | 25 | - |  | - | 0 |  | 0 |
| Lithuania | 1,999 | 297 | - | - | - |  | - | 0 |  | 0 |
| Netherlands | 27,285 | 31,153 | 19,439 | 22,546 | 17,093 |  | 6,289 | 12,157 | 8,725 | 4,925 |
| Norway | 113 | 1,243 | 21 | 12,855 | 526 |  | 7,359 | 129 | 377 | 0 |
| Sweden | 9 | 21 | 36 | 401 | - |  | - | 0 |  | 1 |
| UK (Engl. + | 595 | 6921 | 1,061 | 1,435 | 1,890 |  |  | 935 | 4,401 | 4,198 |
| UK (Scotland) | 300 | 625 | 7 | 4 | 111 |  | 93 | 240 | 172 | 262 |
| Unallocated | -5,004 | -4,960 | 10,869 | 5,964 | -116 |  | 0 | 0 | 0 |  |
| Discard | 78 | 139 | - | 1,036 | 2 |  | 0 | 0 | 0 | 7 |
| Total | 34,240 | 41,105 | 35,705 | 45,881 | 24,144 |  | 7,617 | 21,424 | 18,696 | 13,380 |


|  | Country | 2015 | 2016 |
| :--- | :---: | :---: | :---: |
| Belgium | 63 | 51 | 20171 |
| Denmark | 800 | 268 | 67 |
| Faroe Islands | 0 | 0 | 294 |
| France | 934 | 1,322 | 4 |
| Germany, Fed.Rep. | 644 | 1,879 | 963 |
|  |  |  | 949 |
| Ireland | 0 | 0 | 0 |
| 0Netherlands | 3,305 | 3,892 | 5,638 |
| Norway | 662 | 1,701 | 5 |
| Sweden | 9 | 0 | 0 |
| UK (Engl. + Wales) | 3,581 | 4,697 | 4,546 |
| UK (Scotland) | 0 | 0 | 0 |
| Unallocated | 0 | 0 | 0 |
| Discard | 2,004 | 1,527 | 1,213 |
| Total | 12,002 | 15,337 | 14,579 |

${ }^{1}$ Preliminary

Table 5.6.1. Catches ( $\mathbf{t}$ ) of Trachurus mediterraneus in Divisions 8.ab, 8.c and Sub-Area 7

|  | 7 | 8.AB | 8.c EAST | 8.c West | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 0 | 23 | 3903 |  | 3926 |
| 1990 | 0 | 298 | 2943 |  | 3241 |
| 1991 | 0 | 2122 | 5020 |  | 7142 |
| 1992 | 0 | 1123 | 4804 |  | 5927 |
| 1993 | 0 | 649 | 5576 |  | 6225 |
| 1994 | 0 | 1573 | 3344 |  | 4917 |
| 1995 | 0 | 2271 | 4585 |  | 6856 |
| 1996 | 0 | 1175 | 3443 |  | 4618 |
| 1997 | 0 | 557 | 3264 |  | 3821 |
| 1998 | 0 | 740 | 3755 |  | 4495 |
| 1999 | 0 | 1100 | 1592 |  | 2692 |
| 2000 | 59 | 988 | 808 |  | 1854 |
| 2001 | 1 | 525 | 1293 |  | 1820 |
| 2002 | 1 | 525 | 1198 |  | 1724 |
| 2003 | 0 | 340 | 1699 |  | 2039 |
| 2004 | 0 | 53 | 841 |  | 894 |
| 2005 | 1 | 155 | 1005 |  | 1162 |
| 2006 | 1 | 168 | 794 |  | 963 |
| 2007 | 0 | 126 | 326 |  | 452 |
| 2008 | 0 | 82 | 405 |  | 487 |
| 2009 | 0 | 42 | 1082 |  | 1124 |
| 2010 | 0 | 97 | 370 |  | 467 |
| 2011 | 0 | 119 | 1096 |  | 1225 |
| 2012 | 0 | 186 | 667 | 116 | 969 |
| 2013 | 0 | 52 | 238 | 0 | 290 |
| 2014 | 0 | 130 | 1160 | 0 | 1290 |
| 2015 | 0 | 8 | 890 | 0 | 899 |
| 2016 | 0 | 5 | 471 | 0 | 476 |
| 2017 | 0 | 18 | 684 | 0 | 702 |

Table 5.7.1 Horse mackerel general. Length distributions (\%) Catches by fleet and country in 2017. ( $0 \%=<0.5 \%$ )


Table5.9.1. Summary of the overall sampling intensity on horse mackerel catches in recent years in all areas 1992-2017

| Year | TOTAL CATCH (ICES Estimate) | \% CATCH COVERED bY SAMPLING PROGRAMME* | No. <br> SAMPLES | No. Measured | No. <br> Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 436500 | 45 | 1803 | 158447 | 5797 |
| 1993 | 504190 | 75 | 1178 | 158954 | 7476 |
| 1994 | 447153 | 61 | 1453 | 134269 | 6571 |
| 1995 | 580000 | 48 | 2041 | 177803 | 5885 |
| 1996 | 460200 | 63 | 2498 | 208416 | 4719 |
| 1997 | 518900 | 75 | 2572 | 247207 | 6391 |
| 1998 | 399700 | 62 | 2539 | 245220 | 6416 |
| 1999 | 363033 | 51 | 2158 | 208387 | 7954 |
| 2000 | 247862 | 50 | 378 | 33317 | 4126 |
| 2001 | 257411 | 61 | 467 | 46885 | 7141 |
| 2002 | 223384 | 68 | 540 | 79103 | 6831 |
| 2003 | 223885 | 77 | 434 | 59241 | 8044 |
| 2004 | 195177 | 62 | 518 | 62720 | 9273 |
| 2005 | 212850 | 76 | 573 | 67898 | 8840 |
| 2006 | 190067 | 75 | 602 | 57701 | 9905 |
| 2007 | 164459 | 58 | 397 | 41046 | 8061 |
| 2008 | 179053 | 72 | 488 | 46768 | 8870 |
| 2009 | 229665 | 84 | 902 | 57505 | 10575 |
| 2010 | 227261 | 82 | 710 | 49307 | 14159 |
| 2011 | 221317 | 71 | 502 | 40492 | 7484 |
| 2012 | 191236 | 69 | 501 | 41148 | 8220 |
| 2013 | 183903 | 75 | 686 | 87300 | 9776 |
| 2014 | 149740 | 83 | 650 | 53945 | 8085 |
| 2015 | 110421 | 68 | 825 | 39415 | 7034 |
| 2016 | 114151 | 76 | 1033 | 93853 | 6675 |
| 2017 | 97539 | 63 | 1113 | 116722 | 8221 |

*Percentage related to catch (catch-at-age) acc. to ICES estimation

Table 5.9.2. Horse mackerel sampling intensity for the Western stock in 2017.

| COUNTRY | CATCH | \% CATCH <br> SAMPLED* | NO. <br> SAMPLES | NO. <br> MEASURED | NO. <br> AGED |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Denmark | 4580 | 0 | 0 | 0 | 0 |
| Faroe Islands | 180 | 0 | 0 | 0 | 0 |
| France $^{* *,}{ }^{* * *}$ | 5645 | 0 | 440 | 4383 | 0 |
| Germany $^{\text {Sreland }}$ | 7183 | 68 | 41 | 13730 | 875 |
| Netherlands | 23560 | 95 | 32 | 5782 | 1797 |
| Norway | 14269 | 89 | 50 | 7158 | 1236 |
| Spain | 9885 | 0 | 0 | 0 | 0 |
| Sweden | 16929 | 95 | 960 | 83820 | 3545 |
| UK (England) ${ }^{* * *}$ | 43 |  |  |  |  |
| UK(Scotland) | 612 | 0 | 90 | 624 | 0 |
| Total | 70 | 0 | 53 | 668 | 0 |

*Percentage based on ICES estimate - ** based on length samples from discards in non-targeting horse mackerel fisheries
${ }^{* * *}$ provided only length distributions ${ }^{* * * *}$ based on age sampling

Table 5.9.3. Horse mackerel sampling intensity for the North Sea stock in 2017

| COUNTRY | CATCH | \% CATCH <br> SAMPLED* | NO. <br> SAMPLES | NO. <br> MEASURED | NO. <br> AGED |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Belgium | 67 | 0 | 0 | 0 | 0 |
| Denmark | 340 | 88 | 1 | 111 | 44 |
| Faroe Islands | 4 | 0 | 0 | 0 | 0 |
| France** | 3023 | 0 | 250 | 3118 | 0 |
| Germany | 549 | 0 | 0 | 0 | 0 |
| Netherlands | 5637 | 93 | 29 | 6121 | 724 |
| Norway | 0 | 0 | 0 | 0 |  |
| UK (England) | 14579 | 46 | 0 | 0 | 0 |
| Total |  |  |  | 90350 | 768 |

*Percentage based on ICES estimate. ** provided only length distributions
*** based on age sampling

### 5.12 Figures



Figure 5.1.1a. Horse mackerel catches ${ }^{15 t}$ quarter 2017.


Figure 5.1.1b. Horse mackerel catches $2^{\text {nd }}$ quarter 2017.


Figure 5.1.1c. Horse mackerel catches $3^{\text {rd }}$ quarter 2017.


Figure 5.1.1d. Horse mackerel catches $4^{\text {th }}$ quarter 2017.


Figure 5.2.1: Distribution of Horse Mackerel in the Northeast-Atlantic: Stock definitions as used by the 2004 WG MHSA. Note that the "Juvenile Area" is currently only defined for the Western Stock distribution area - juveniles do also occur in other areas (like in Div. 7.d). Map source: GEBCO, polar projection, 200 m depth contour drawn.


Figure 5.3.1. Total catch for Western Horse Mackerel stock, period 1982-2017.


Figure 5.3.4. Total catch for North Sea Horse Mackerel stock, period 1982-2017

Horse mackerel Stocks 1982-2017 Catches by stock


Figure 5.4.1 Horse mackerel general. Total catches in the northeast Atlantic during the period 19822017. The catches taken from the southern, western and North Sea horse mackerel stocks are shown in relation to the total catches in the northeast Atlantic. Catches from Div. 8.c were transferred from southern stock to western stock from 1982 onwards. Southern horse mackerel is assessed by ICES WGHANSA since 2011.


Figure 5.4.2. North Sea horse mackerel stock. Total catches by Division during the period 1982-2017.


Figure 5.4.3. Western horse mackerel stock. Total catches by Sub-Area during the period 1982-2017.


Figure 5.9.1 North Sea horse mackerel stock. Percentage sampled catch (blue) vs. unsampled catch (red) by division and year. Period 2000-2017.


Figure 5.9.2. North Sea horse mackerel stock. Sampling intensity index as percentage sampled catch in total catch by year. Period 2000-2017.


Figure 5.9.5. Western horse mackerel stock. Percentage sampled catch (blue) vs. unsampled catch (red) by division and year. Period 2000-2017. Area of distribution of Western stock was divided into different regions. Chan: (7.e,f,h); W- SCO+IRL (7.a-c, 7.j-k and 6.a); BoB (8.a,b,d); CanSea(8.c); NNsea (3.a and 4.a); NOR (2.a and 5.a).


Figure 9.5.6. Western horse mackerel stock. Sampling intensity index as percentage sampled catch in total catch by year. Period 2000-2017.

## 6 Horse Mackerel: Divisions 27.4.a (Q1 and Q2), 27.3.a (excluding Western Skagerrak Q3 and Q4), 27.4.b, 27.4.c and 27.7.d

### 6.1 ICES Advice Applicable to 2018

In 2012 the North Sea horse mackerel (NSHM) was classified as a category 5 stock, based on the ICES approach to data-limited stocks (DLS). Since then, a progressive reduction of TAC was advised by ICES, from 25500 tonnes in 2013-2014 to 15200 tonnes in 2015-2016. This reduction in the advised catch was supported by the analysis of information from the North Sea International Bottom Trawl Survey (NS-IBTS) traditionally used in the assessment, but also new information from the Channel Ground Fish Survey (CGFS) since 2014. Despite the considerable increase showed by the CGFS in 2015 survey, due to the high uncertainty in the two survey indices, catch in 2017 was advised to continue at 15200 tonnes. However, new information indicated a $16.7 \%$ discards of NSHM in 2015 in non-directed fisheries. This new information is taken into account in the catch advice for 2017. The advice landings were 15200 tonnes and the advice total catch was 18247 tonnes.

In 2017 this stock was benchmarked and the NS-IBTS and CGFS survey indices where modelled together. The resulting joint index was considered a proper indication of trend in abundance over time and the NSHM stock was upgraded to category 3. The joint survey index showed in 2016 a continuation of the increasing trend started in 2013. The application of the HCR 3.1 (ICES, 2012, comparison of the two latest index values with the three preceding values multiplied by the recent advised catch) resulted in an increase higher than $20 \%$ in the catch advice for 2018 in comparison to advice for 2017. Accordingly the uncertainty cap was applied. In addition, Length Based DLS methods indicated that the F was in 2016 slightly above the FMSY proxy, and stock size relative to reference points was unknown. Therefore, the precautionary buffer was also applied to the advice, since it hadn't been applied since 2014. This resulted in a catch advice for 2018 and 2019 no more than 17517 tonnes. Considering the $13.35 \%$ average discards, were advice not being higher than 15179 tonnes.

### 6.2 Fishery of North Sea horse mackerel stock

Based on historical catches taken by the Danish industrial fleet for reduction into fishmeal and fish oil in the 1970s and 1980s, approximately $48 \%$ of the EU North Sea horse mackerel TAC was taken by Denmark. Catches were taken in the fourth quarter mainly in Divisions 27.4.b and 27.7.d. The 1990s saw a drop in the value of industrial fish, limited fishing opportunities and steep increases in fuel costs that affected the Danish quota uptake. In 2001, individual quota scheme for a number of species was introduced in Denmark, but not for North Sea horse mackerel. This lead to a rapid restructuring and lower capacity of the Danish fleet, which in combination with the above mentioned factors led to a decrease of the Danish North Sea horse mackerel catches.

Since the 1990's, a larger portion of catches has been taken in a directed horse mackerel fishery for human consumption by the Dutch freezer-trawler fleet. This is possible because Denmark has traded parts of its quota with the Netherlands for other species. However due to the structure of the Danish quota management setup only a limited amount of quota can be made available for swaps with other countries. These practical implications of the management scheme largely explain the consistent underutilisation of the TAC over the period 2010-2013 (approximately 50\%). However, following the sharp reduction in TAC in 2014 uptake increased significantly to above $80 \%$ in 2015
and $100 \%$ in 2016 (see Figure 6.2.1), although an important part of these catches were discards ( $16.7 \%$ and $10 \%$ respectively). In 2017 the $80 \%$ of the TAC was used, with an 8.3\% discards.

Catches taken in Divisions 27.3.a and 27.4.a during the two first quarters and all year in Divisions 27.4.b, 4.c and 27.7.d are regarded North Sea horse mackerel (Section 5, Table 5.4.1). The catches were relatively low during the period 1982-1997 with an average of 18000 tonnes, but increased between 1998 (30500 tonnes) and 2000 (45130 tonnes). From 2000 to 2010, the catches varied between 24149 and 45883 tonnes. Since 2014 a steep decline in catches is observed, both due to the reduction in the TAC since 2014 but also the underutilization of the quota. In 2017 the catch was 14579 tonnes, with an $82 \%$ of total catch being caught in area 27.7.d.

Over the period 1985-2001 most catches were taken in the area 27.4.b (Figure 6.2.3). However, since 2002 the proportion of catches from area 27.7.d increased steadily until 2013, when the $92 \%$ of total catches were fished in this area (Figure 6.2.2). Germany, UK and Netherlands accounted for most of the landings, that were taken in quarter 1, but especially in quarter 4 (Figure 6.2.3). Most of the discards were reported in area 7d, more importantly during quarters $3^{\text {rd }}$ and $4^{\text {th }}$, by the French bottom-trawl fleet. Discarding in the target pelagic fisheries is considered negligible. New information in 2015 from bottom-trawl fisheries not directed at horse mackerel indicated an overall discard rate of $16.7 \%$ for the stock as a whole, while in 2016 this rate is $10 \%$. Complete discard information for earlier years has not been submitted to ICES. However, information from national discard reports for the non-directed bottom-trawl fisheries indicates a similar level of discarding in earlier years.

### 6.3 Biological Data

### 6.3.1 Catch in Numbers at Age

In 2017, as already occurred in 2016, there has been a marked reduction in the coverage of biological sampling. Only the $38 \%$ of landings was sampled, in comparison to 2013 and 2014 when $71 \%$ and $63 \%$ were sampled respectively (section 5 figure 5.9.1). In addition, this low coverage was carried out mostly by the Dutch fleet in quarter Q4 and divisions 27.7.d and 27.4.c. Despite most landing catch was taken from this area and quarter ( $81.9 \%$ of landings in division 27.7.d and $75 \%$ in quarter Q4, Figure 6.3.1) still part of landings were fished in other areas and quarters. In order to avoid a biased perception of the age distribution of catches over the year and areas, this partial and uneven sampling effort should be avoided in future years.

Annual catch numbers at age by area for year 2017 are shown in table 6.3.1. Due to the low level of sampling effort out of area 27.7.d., there is not enough information to represent age distribution in those areas, and hence, the one observed in 27.7.d is taken as the basis to separate catch by age. Catch-at-age for the whole period 1995-2017 are given in, table 6.3.2 and in Figure 6.3.2. These data show that since 2005 the age distribution of catches has experienced a reduction, with a decrease in the range of ages of importance in total catches. In parallel to the rejuvenation of catches, the comparison of catch-at-age data after 1998 by area (Figure 6.3.3) shows that since 2010 commercial catches have increased in area 27.7.d in comparison to the areas 27.3.a and $4 \mathrm{a}, \mathrm{b}$ and c where the opposite pattern was found. Since 2015, commercial catches are focused mostly on cohort 2014 that was the main component of catches both in and out of the area 27.7.d at age 1 in year 2015, age 2 in 2016 and age 3 in 2017. Ages 1 and 2 appear with moderate importance in the total catch.

Although 2015 cohort seems to be clear in the catch-at-age distribution, in general cohort structure is not clearly detectible in the data. This may partly be due to the shifts in distribution of the fishery. In addition, it may partly be due to age reading difficulties, which are a known to be encountered (e.g. Bolle et al., 2011). Most clearly detectable is the relatively large 2001 year class, although it is not clearly present in the catch in all years. There are indications that environmental circumstance may be an important factor (possibly stronger than stock size) contributing to spawning success in horse mackerel. This is for example illustrated by the largest year classes (1982 and 2001) observed in the Western stock which incidentally were produced at the lowest observed stock sizes. Since 2001 is considered to have been a relatively strong year class in the Western stock as well, it is plausible that circumstances in the North Sea were similar to those in Western areas and also allowed for relatively high spawning success in the North Sea.

Lastly, potential mixing of fish from the Western and North Sea stock in area 27.7.d and 27.7.e in winter may also confuse the cohort signals. For example, the large recruitment in the Western stock may have led to more of these fish being located in the North Sea stock area as age 1 fish in 2002. With the intention of clarifying the mixing among the North Sea and the Western horse mackerel stocks, and how this may affect to the age distribution of catches. In 2015 was conducted by IMARES and the Pelagic Freezertrawler Association (PFA). The results of this project were not conclusive because of the low sample size and some technical glitches in the sequencing. The chemical analysis generated some new insights but also some more questions on the variability that could be expected between years and seasons. Currently more genetic samples are being taking by PFA in different areas of the North East Atlantic. In addition, catch sampling carried out by several pelagic fishery companies is being explored to give information on the separation between North Sea and Western horse mackerel. Until the mixing of both stocks is clarified and catch-at-age data can be clearly segregated the development of analytical assessment will be limited.

### 6.3.2 Mean weight at age and mean length at age

The mean weight and mean length-at-age in the commercial catches of 2017 are presented in tables 6.3.3 and 6.3.4 respectively by quarter and area. As explained for the distribution of catch-at-age by area, due to the biased sampling coverage in 2017 for several ages mean weight and length in quarters Q2-Q3 in areas 27.3.a, 24.7.a-b-c are assumed the same than in quarter $\mathrm{Q} 1-\mathrm{Q} 4$ in area 27.7.d.

The mean annual weight and length over the period 2000-2016 are presented in table 6.3.2 and figures 6.3.4 and 6.3.5 respectively. Despite there are no strong differences over this period, since 2006 there seems to be a slight but steady increase in both weight and length until 2015, when a declining pattern is observed. It may be hypothesized that this is due to density-dependent effects, due to the relatively successful recruitment of 2015.

### 6.3.3 Maturity-at-age

Peak spawning in the North Sea falls in May and June (Macer, 1974), and spawning occurs in the coastal regions of the southern North Sea along the coasts of Belgium, the Netherlands, Germany, and Denmark.

There is no information available about the maturity-at-age of the North Sea Horse mackerel stock.

### 6.3.4 Natural mortality

There is no specific information available about natural mortality of this stock.

### 6.4 Data Exploration

### 6.4.1 Catch curves

The log-catch numbers were plotted by cohort to estimate the negative gradient of the slope and get an estimate of total mortality (Z). Fully selected ages 3 to 14 from the 1992 - 2016 period provide complete data for the 1992 to 2006 cohorts (Figure 6.4.1). The estimated negative gradients by cohort (Figure 6.4.2) indicate an increasing trend in total mortality for the period examined, with a marked increment in the cohorts 2005 and 2006. However, due to the low quality of the signals for some cohorts these Z estimates has to be considered with caution.

An analysis of the catch number at age data carried out in 2011 showed that only the $1 \mathrm{vs} .2,2 \mathrm{vs} .3,7 \mathrm{vs} .8$ and 8 vs .9 age groups were positively and significantly correlated in the catch. This analysis was not updated this year but these results suggest limitations in the catch-at-age data.

### 6.4.2 Assessment models and alternative methods to estimate the biomass

In 2002 Ruckert et al. estimated the North Sea horse mackerel biomass based on a ratio estimate that related cpue data from the IBTS to cpue data of whiting (Merlangius merlangus). The applied method assumes that length specific catchability of whiting and horse mackerel are the same for the IBTS gear. Subsequently, they use the total biomass of whiting derived from an analytical stock assessment (MSVPA) to estimate the relationship between cpue and biomass.

At the 2014 WGWIDE some exploratory model fits were attempted with the JAXass model, using the data available. The JAXass (JAX assessment) model is a simple statistical catch-at-age model fitted to an age-aggregated index of ( $2+$ ) biomass, total catch data and proportions at age from the catch. It is based on Per Sparre's "separable VPA" model, an ad hoc method tested for the first time at WGWIDE in 2003, and later 2004. A new analysis using this model was also done in 2007 using an IBTS index. In 2014 the model has been coded in ADMB (Fournier et al., 2012) and updated with an improved objective function (dnorm), extra years of data and new methods for calculating the index (see above).
Difficulties in fitting an assessment model for this stock include:

- Unclear stock boundaries
- Difficulty aging horse mackerel
- Lack of strong cohort signals in CAA data.
- Scientific index derived from a survey not specifically designed for horse mackerel and not covering one of the main fishing grounds for the stock (VIId)

Catches taken in area 27.7. d are close to the management boundary between the (larger) western horse mackerel stock and the NS horse mackerel stock. It is quite possible that given changes in oceanographic condition, or changes in abundance of either of the two stocks, that some proportion of the catches taken in area 27.7.d actually originated from the western horse mackerel stock. Nevertheless, all assessment models
used in the MSE assume that $100 \%$ of fish caught in area 27.7.d belong to the North Sea horse mackerel stock. This is in agreement with stock and management definitions.

### 6.4.3 Survey data

### 6.4.3.1 Egg Surveys

No egg surveys for horse mackerel have been carried out in the North Sea since 1991. Such surveys were carried out during the period 1988-1991. SSB estimates are available historically. However, they were calculated assuming horse mackerel to be a determinate spawner. Horse mackerel is now considered an indeterminate spawner. Therefore egg abundance could only be considered a relative index of SSB. The mackerel egg surveys in the North Sea do not cover the spawning area of horse mackerel.

### 6.4.3.2 IBTS Survey Data

Many pelagic species are frequently found close to the bottom during daytime (which is when the IBTS survey operates) and migrate upwards predominantly during the night they are susceptible to semi-pelagic fishing gear and to bottom trawls (Barange et al. 1998). Eaton et al. (1983) argued that horse mackerel of 2 years and older are predominantly demersal in habit. Therefore, in the absence of a targeted survey for this stock, the IBTS is considered a reasonable alternative. IBTS data are also used in the assessment of the southern horse mackerel stock.

IBTS data from quarter Q3 were obtained from DATRAS and analysed. Based on a comparison of IBTS data from all 4 quarters in the period 1991-1996, Ruckert et al. (2002) showed that horse mackerel catches in the IBTS were most abundant in the third quarter of the year. In 2013 WGWIDE considered that using an 'exploitable biomass index' estimated with the abundance by haul of individuals larger than 20 cm is the most appropriate for the purpose of interpreting trend in the stock.

To create indices, a subset of ICES rectangles was selected. Rectangles that were not covered by the survey more than once during the period 1991-2012 were excluded from the index area. In 2012, WGWIDE expressed concern that the previously selected index area did not sufficiently cover the distribution area of the stock, especially in years that the stock would be relatively more abundant and spread out more. Ruckert et al. (2002) also identified a larger distribution area of the North Sea stock. Based on the above, 61 rectangles were identified to be included in the index area as shown in Figure 6.4.3.

### 6.4.3.3 The French Channel Groundfish Survey (CGFS) in Q4

In order to improve data basis for the North Sea horse mackerel assessment, alternative survey indices have been explored. Previous indices used had only cover the North Sea distribution of the stock, while the majority of catches in recent years have come from the eastern English Channel (27.7.d). We evaluated the potential contribution of the French Channel Ground Fish Survey in 27.7.d (CGFS) in Quarter 4. The CGFS is carried out since 1990 and has frequent captures of horse mackerel. Though this survey is conducted in a different quarter to the North Sea IBTS, the observed seasonal migration patterns of horse mackerel indicate that fish move into the channel following quarter Q3, so the timing is considered appropriate.

In 2015, the RV "Gwen Drez" was replaced by the RV "Thalassa" to carry out the CGFS survey. In 2014 an inter-calibration process was conducted to quantify the differences
in catchability for a large number of species. ICES reviewed this inter-calibration exercise and found a number of drawbacks that may undermine the reliability of the estimated conversion factors. The main concerns were:

- The analyses were limited in the number of tows. Considering that a number of these tows could be zeros for one of the two vessels and possibly resulting in highly uncertain estimates.
- Lack of length-specific correction factor.
- At a standardized depth of 50 m and above, wing spread estimates for the R/V Thalassa as measured by the MARPORT sensor were deemed erroneous, which may question the validity of estimated area swept by the net on the R/V Thalassa and the effect it may have on correction factors for species caught at depth at 50 m and greater.
- A number of tow locations including areas outside 27.7.d were excluded. Changing the depth range of a survey can add serious bias in the calibration and the current approach seems to be ignoring this issue.
- Correction coefficients were not measured without error.

However, these limitations were considered by WGWIDE to be of minor importance for the North Sea horse mackerel since:

- Despite being still a low sample size the North Sea horse mackerel was present in all the 32 paired hauls.
- There are no important differences in size distribution (Figure 6.4.4).
- The analysis with and without the areas excluded in the new sampling design did not show important differences (ICES, 2017).
- cpue or North Sea horse mackerel for hauls deeper than 50 m was relatively low (Figure 6.4.5), and it is expected than the potential problems in determining the conversion factor bellow that depth range would have a relatively minor impact in the estimated abundance.

For these reasons it was finally decided to continue using the CGFS survey, standardizing the time-series of abundance for the period 1990-2015 with the estimated conversion factor 10.363.

### 6.4.3.4 Norht Sea horse mackerel benchmark exercise

In January 2017, a benchmark process was conducted for NSHM (ICES, 2017). Based on capacity to model the overdispersion and the high proportion of zero values in the survey catch data the hurdle models was concluded the best option of all the model alternatives tested. The log-likelihood ratio test, the AIC and the evidence ratio statistic supported that the model that best represented the data was a hurdle model with Year and Survey as explanatory factors (including the interaction term) in the count model (GLM-negative binomial), and Year and Survey (without the interaction) in the zero model (GLM-binomial).

The probability of having a cpue zero was modelled by a logistic regression with a GLM-binomial distribution model:

$$
\operatorname{logit}\left(\pi_{i}\right)=\text { Intercept }_{z e r o}+\text { Year }_{i, z e r o}+\text { Survey }_{i, \text { zero }}
$$

Where $\pi_{i}$ is the mean probability of having a cpue zero as a function Year and Survey.

The expected cpue of North Sea horse mackerel, conditional to not having a zero in hurdle models (not having a false zero in zero inflated models) was modelled with a GLM-negative binomial distribution model:

$$
\log \left(C P U E_{i}\right)=\text { Intercept }_{\text {count }}+\text { Year }_{i, \text { count }} x \text { Survey }_{i, \text { count }}
$$

This model was used to synthesise the information from both the GCFS and IBTS and predict the average annual cpue index per haul as an indicator of trends in stock abundance both for the juvenile $(<20 \mathrm{~cm})$ and exploitable $(>20 \mathrm{~cm})$ substocks. The contribution of the two surveys to the combined index is weighted taken into consideration their respective surface coverage as well as the mean wing spread. This index model allowed upgrading the NSHM to a category 3 stock within the ICES classification.

### 6.4.4 Summary of index trends

The survey index for both the small and exploitable substocks experienced a marked decline in the early-mid 2000s (Figure 6.4.6; table 6.4.1). This reduction was due in part to the decline of the average abundance by haul over time, but also to the increase of hauls with zero catch of horse mackerel, from $26 \%$ in 1998 to the highest observed value of $72 \%$ in 2013 (Figure 6.4.7). Since 2014 a slight decrease in zero hauls was observed in juveniles group (smaller than 20 cm ). Since 2013, in addition to the decline of zero hauls, the mean cpue in the non-zero hauls has increased. After an increase in 2016, the abundance survey index for the exploitable substock has shown a marked decline in 2017. This pattern has been mostly due to the decline of the survey index estimated for the CGFS in comparison to the value in 2016 (Figure 6.4.8). The survey index of the juvenile substock, that also showed an increasing pattern since 2013, seems to be stabilized since 2014 in the CGFS, but in the IBTS is in 2017 at the lowest level since 1992. Due to this compensation by the CGFS, the abundance index for juveniles, show a less steep decline than the exploitable substock index. The size distribution in both the CGFS and the IBTS suggest the entrance of a moderate new cohort in 2017 (between 47 cm in the IBTS and $7-11 \mathrm{~cm}$ in the CGFS) age 0 (Figures 6.4.9 and 6.4.10).

However, despite the index of abundance of individuals smaller than 20 cm could be considered a recruitment index, it has to be considered with caution. Preliminary examinations of how the juvenile $(0-19 \mathrm{~cm})$ indices relate to subsequent exploitable abundance $(20+\mathrm{cm})$ do not indicate strong linkages. The very high juvenile indices in the early 2000s in the IBTS were not subsequently picked up in the exploitable component. Hence while increases in the juvenile indices are encouraging, whether these lead to increases in the exploitable component of the stock need to be confirmed in the future with observations in the $20+\mathrm{cm}$ indices.

### 6.4.5 Data Limited Stock methods and MSY proxy reference points

As part of the ICES approach to provide advice within the MSY framework for stocks of category 3 and 4, different Data Limited methods to estimate MSY proxy reference points for the North Sea horse mackerel were explored. This analysis and results were presented in a separate working document (Pérez-Rodríguez, 2017). After exploring the compliance with each method assumptions and assessing the data availability the group decided that the Length Based Indicators would be the only DLS method to be applied to the NSHM.

Despite this length based method will have to be applied in the future to a longer timeseries of catch length frequencies, only length data have been collected for 2016 and 2017. The estimates of F/Fmsy proxy indicate that fishing mortality seems to be, both in

2016 and 2017, slightly above FMSY for the North Sea horse mackerel, with F/FMSY=1.082 and 1.073 respectively (Figure 6.4.11).

### 6.4.6 Ongoing work

To improve the knowledge base for North Sea horse mackerel about the degree of connection and migrations in between the North Sea and the Western Stock, catch sampling carried out by several pelagic fishery companies is being explored to give information on the separation between North Sea and Western horse mackerel. To improve the abundance indicators the potential application of a commercial fishery search time index will be explored. Horse mackerel is fished while it is very close to the bottom in relatively dispersed, small schools. The fishery is mostly executed using long hauls and there may be extensive search time involved. Handled in an appropriate statistical framework, taking into account the nature of the fishery and other factors such as seasonality and alternative fishing opportunities, the search time and catch rates could provide for an indication of changes in stock size over time. Catch rates in areas 27.7.e, 27.7.d and southern North Sea will be analysed from skippers' private logbooks.

### 6.5 Basis for 2019 Advice. ICES DLS approach.

Stock advice for NSHM is biannual. In 2017 the advice for years 2018 and 2019 was provided. The joint abundance survey index indicated a continuation in the increasing trend observed since 2013. This increase was due mostly to the increment observed in the CGFS survey index. Despite that, as mentioned in the previous section, the joint survey index for 2017 has shown a sudden change and steep decline due to the drop of the CGFS survey index, WGWIDE decided to continue with the current advice of 17517 tonnes for 2019.

The fisheries in the area have largely been focused on the smaller fish in 2015, 2016, 2017 and it is expected that this will continue in 2018 and 2019. With this pattern of exploitation, mostly immature individuals are caught which might hinder the recovery of the stock by removing an important portion of the recent year classes before they enter the spawning stock.

### 6.6 Management considerations

In the past, Division 27.7.d was included in the management area for Western horse mackerel together with Divisions 27.2.a, 27.7.a-c, 27.7.e-k, 27.8.a, 27.8.b, 27.8.d, 27.8.e, Subarea 6, EU and international waters of Division 5.b, and international waters of Subareas 12 and 14. ICES considers Division 27.7.d to be part of the North Sea horse mackerel distribution area. Since 2010, the TAC for the North Sea area has included Divisions 27.4.bc and 27.7.d. Considering that a majority of the catches are taken in Division 27.7.d, the total of North Sea horse mackerel catches are effectively constrained by the TAC since the realignment of the management areas in 2010.

Catches in Divisions 27.3.a (Western Skagerrak) and 27.4.a in quarters 3 and 4 are considered to be from the Western horse mackerel stock, while catches in quarters 1 and 2 are considered to be from the North Sea horse mackerel stock. Catches in area 27.4.a and 27.3.a are variable. In recent years only Norway has had significant catches in this area, but these are only taken in some years.

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### 6.8 Tables

Table 6.3.1. North Sea Horse Mackerel stock. Catch in numbers (1000) by quarter and area in 2017.

| Number/1000 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.99 | 44.35 | 96.06 | 10.51 | 0 | 151.91 |
| 2 | 0.9 | 40.45 | 87.61 | 9.59 | 1579.74 | 1718.29 |
| 3 | 2.73 | 122.35 | 265.02 | 29 | 6243.76 | 6662.86 |
| 4 | 0.4 | 17.94 | 38.85 | 4.25 | 4502.28 | 4563.72 |
| 5 | 0.25 | 11.08 | 24 | 2.63 | 3308.95 | 3346.91 |
| 6 | 0.15 | 6.87 | 14.89 | 1.63 | 1229.79 | 1253.33 |
| 7 | 0.11 | 4.99 | 10.8 | 1.18 | 2074.86 | 2091.93 |
| 8 | 0.1 | 4.45 | 9.63 | 1.05 | 1604.79 | 1620.02 |
| 9 | 0.04 | 1.75 | 3.78 | 0.41 | 558.59 | 564.58 |
| 10 | 0.01 | 0.49 | 1.06 | 0.12 | 261.55 | 263.23 |
| 11 | 0.02 | 1.02 | 2.2 | 0.24 | 297.05 | 300.53 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.01 | 0.49 | 1.06 | 0.12 | 261.55 | 263.23 |
| 15 | 0.01 | 0.49 | 1.06 | 0.12 | 261.55 | 263.23 |
| Sum | 5.72 | 256.7 | 556.04 | 60.84 | 22184.47 | 23063.76 |
| 2Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 16.65 | 24.39 | 17.35 | 411.9 | 828.12 | 1298.41 |
| 2 | 15.18 | 22.24 | 15.83 | 375.67 | 755.27 | 1184.19 |
| 3 | 45.93 | 67.28 | 47.87 | 1136.4 | 2284.7 | 3582.19 |
| 4 | 6.73 | 9.86 | 7.02 | 166.61 | 334.96 | 525.18 |
| 5 | 4.16 | 6.09 | 4.34 | 102.93 | 206.94 | 324.46 |
| 6 | 2.58 | 3.78 | 2.69 | 63.83 | 128.32 | 201.19 |
| 7 | 1.87 | 2.74 | 1.95 | 46.31 | 93.1 | 145.97 |
| 8 | 1.67 | 2.44 | 1.74 | 41.29 | 83.02 | 130.17 |
| 9 | 0.66 | 0.96 | 0.68 | 16.22 | 32.61 | 51.13 |
| 10 | 0.18 | 0.27 | 0.19 | 4.55 | 9.15 | 14.35 |
| 11 | 0.38 | 0.56 | 0.4 | 9.45 | 18.99 | 29.77 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |


| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14 | 0.18 | 0.27 | 0.19 | 4.55 | 9.15 | 14.35 |
| 15 | 0.18 | 0.27 | 0.19 | 4.55 | 9.15 | 14.35 |
| Sum | 96.37 | 141.15 | 100.44 | 2384.26 | 4793.5 | 7515.73 |


| 3Q |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 59.57 | 59.57 |
| 1 | 0 | 0 | 70.95 | 111.67 | 1329.97 | 1512.59 |
| 2 | 0 | 0 | 64.71 | 101.85 | 1212.97 | 1379.53 |
| 3 | 0 | 0 | 195.75 | 308.08 | 3669.25 | 4173.08 |
| 4 | 0 | 0 | 28.7 | 45.17 | 537.94 | 611.81 |
| 5 | 0 | 0 | 17.73 | 27.9 | 332.34 | 377.98 |
| 6 | 0 | 0 | 10.99 | 17.3 | 206.08 | 234.38 |
| 7 | 0 | 0 | 7.98 | 12.55 | 149.52 | 170.05 |
| 8 | 0 | 0 | 7.11 | 11.19 | 133.33 | 151.64 |
| 9 | 0 | 0 | 2.79 | 4.4 | 52.38 | 59.57 |
| 10 | 0 | 0 | 0.78 | 1.23 | 14.7 | 16.72 |
| 11 | 0 | 0 | 1.63 | 2.56 | 30.5 | 34.68 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0.78 | 1.23 | 14.7 | 16.72 |
| 15 | 0 | 0 | 0.78 | 1.23 | 14.7 | 16.72 |
| Sum | 0 | 0 | 410.69 | 646.39 | 7757.96 | 8815.04 |
| 4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 391.05 | 391.05 |
| 1 | 0 | 0 | 6.18 | 1874.01 | 15901.14 | 17781.33 |
| 2 | 0 | 0 | 5.64 | 4425.6 | 12187.97 | 16619.21 |
| 3 | 0 | 0 | 17.06 | 11524.87 | 36654.11 | 48196.04 |
| 4 | 0 | 0 | 2.5 | 1611.53 | 2872.44 | 4486.47 |
| 5 | 0 | 0 | 1.55 | 250.28 | 1730.29 | 1982.11 |
| 6 | 0 | 0 | 0.96 | 0 | 1722.52 | 1723.48 |
| 7 | 0 | 0 | 0.7 | 0 | 413.71 | 414.4 |
| 8 | 0 | 0 | 0.62 | 0 | 542.37 | 542.99 |
| 9 | 0 | 0 | 0.24 | 0 | 263.86 | 264.1 |
| 10 | 0 | 0 | 0.07 | 0 | 0 | 0.07 |
| 11 | 0 | 0 | 0.14 | 0 | 173.56 | 173.7 |


| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0.07 | 0 | 0 | 0.07 |
| 15 | 0 | 0 | 0.07 | 0 | 0 | 0.07 |
|  | 0 | 0 | 786.34 | 999.45 | 80108.12 | 81893.91 |
| 1-4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 450.61 | 450.61 |
| 1 | 17.64 | 68.73 | 190.55 | 2408.1 | 18059.23 | 20744.25 |
| 2 | 16.09 | 62.69 | 173.79 | 4912.7 | 15735.95 | 20901.21 |
| 3 | 48.66 | 189.63 | 525.7 | 12998.35 | 48851.82 | 62614.16 |
| 4 | 7.13 | 27.8 | 77.07 | 1827.55 | 8247.61 | 10187.17 |
| 5 | 4.41 | 17.18 | 47.62 | 383.74 | 5578.52 | 6031.46 |
| 6 | 2.73 | 10.65 | 29.53 | 82.76 | 3286.72 | 3412.39 |
| 7 | 1.98 | 7.73 | 21.42 | 60.04 | 2731.19 | 2822.36 |
| 8 | 1.77 | 6.89 | 19.1 | 53.54 | 2363.51 | 2444.81 |
| 9 | 0.69 | 2.71 | 7.5 | 21.03 | 907.44 | 939.38 |
| 10 | 0.19 | 0.76 | 2.11 | 5.9 | 285.41 | 294.37 |
| 11 | 0.4 | 1.58 | 4.37 | 12.25 | 520.09 | 538.69 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.19 | 0.76 | 2.11 | 5.9 | 285.41 | 294.37 |
| 15 | 0.19 | 0.76 | 2.11 | 5.9 | 285.41 | 294.37 |
| Sum | 102.09 | 397.85 | 1102.97 | 22777.77 | 107588.92 | 131969.6 |

Table 6.3.2. Numbers at age (millions), weight at age (kg) and length at age (cm) for the North Sea horse mackerel 1995-2017 in the commercial fleet catches.


| KG | WEIGHT |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| 1 | 0.076 | 0.107 | 0.063 | 0.063 | 0.063 | 0.075 | 0.067 | 0.066 | 0.075 | 0.076 | 0.07 | 0.074 | 0.615 | 0.063 | 0.074 | 0.077 | 0.061 | 0.069 | 0.077 | 0.078 | 0.062 | 0.07 | 0.06 |
| 2 | 0.126 | 0.123 | 0.102 | 0.102 | 0.102 | 0.1 | 0.09 | 0.096 | 0.105 | 0.105 | 0.087 | 0.098 | 0.081 | 0.096 | 0.087 | 0.101 | 0.092 | 0.09 | 0.099 | 0.11 | 0.099 | 0.093 | 0.086 |
| 3 | 0.125 | 0.143 | 0.126 | 0.126 | 0.126 | 0.137 | 0.094 | 0.129 | 0.122 | 0.122 | 0.104 | 0.116 | 0.104 | 0.109 | 0.113 | 0.118 | 0.096 | 0.118 | 0.112 | 0.113 | 0.13 | 0.115 | 0.113 |
| 4 | 0.133 | 0.156 | 0.142 | 0.142 | 0.142 | 0.152 | 0.117 | 0.155 | 0.136 | 0.146 | 0.133 | 0.124 | 0.115 | 0.125 | 0.134 | 0.137 | 0.115 | 0.142 | 0.138 | 0.135 | 0.15 | 0.126 | 0.131 |
| 5 | 0.146 | 0.177 | 0.16 | 0.16 | 0.16 | 0.165 | 0.159 | 0.171 | 0.164 | 0.174 | 0.159 | 0.141 | 0.13 | 0.145 | 0.152 | 0.155 | 0.145 | 0.152 | 0.166 | 0.144 | 0.169 | 0.158 | 0.173 |
| 6 | 0.164 | 0.187 | 0.175 | 0.175 | 0.175 | 0.192 | 0.183 | 0.195 | 0.18 | 0.198 | 0.197 | 0.178 | 0.163 | 0.161 | 0.182 | 0.183 | 0.166 | 0.172 | 0.18 | 0.177 | 0.196 | 0.155 | 0.189 |
| 7 | 0.161 | 0.203 | 0.199 | 0.199 | 0.199 | 0.194 | 0.198 | 0.216 | 0.193 | 0.224 | 0.238 | 0.212 | 0.192 | 0.193 | 0.195 | 0.206 | 0.193 | 0.183 | 0.2 | 0.184 | 0.26 | 0.162 | 0.177 |
| 8 | 0.178 | 0.195 | 0.231 | 0.231 | 0.231 | 0.216 | 0.201 | 0.227 | 0.212 | 0.229 | 0.248 | 0.247 | 0.197 | 0.221 | 0.258 | 0.199 | 0.193 | 0.188 | 0.216 | 0.201 | 0.29 | 0.235 | 0.188 |
| 9 | 0.165 | 0.218 | 0.25 | 0.25 | 0.25 | 0.244 | 0.237 | 0.228 | 0.24 | 0.256 | 0.259 | 0.236 | 0.257 | 0.286 | 0.253 | 0.241 | 0.305 | 0.212 | 0.223 | 0.222 | 0.265 | 0.246 | 0.222 |
| 10 | 0.173 | 0.241 | 0.259 | 0.259 | 0.259 | 0.283 | 0.246 | 0.253 | 0.27 | 0.29 | 0.287 | 0.286 | 0.255 | 0.295 | 0.322 | 0.227 | 0.334 | 0.204 | 0.226 | 0.22 | 0.312 | 0.359 | 0.233 |
| 11 | 0.317 | 0.307 | 0.3 | 0.3 | 0.3 | 0.286 | 0.26 | 0.303 | 0.24 | 0.3 | 0.335 | 0.237 | 0.517 | 0.273 | 0.422 | 0.284 | 0.345 | 0.275 | 0.242 | 0.264 | 0.262 | 0.369 | 0.257 |
| 12 | 0.233 | 0.211 | 0.329 | 0.329 | 0.329 | 0.354 | 0.286 | 0.293 | 0.298 | 0.297 | 0.349 | 0.261 | 0.279 | 0.309 | 0.447 | 0.234 | 0.408 | 0.195 | 0.263 | 0.287 | 0.318 | 0.379 |  |
| 13 | 0.241 | 0.258 | 0.367 | 0.367 | 0.367 | 0.316 | 0.287 | 0.317 | 0.356 | 0.301 | 0.338 | 0.267 | 0.339 | 0.375 | 0.383 | 0.288 | 0.474 |  | 0.262 | 0.252 | 0.351 | 0.242 |  |
| 14 | 0.348 | 0.277 | 0.299 | 0.299 | 0.299 |  | 0.295 | 0.32 | 0.316 | 0.338 | 0.373 | 0.302 | 0.414 | 0.277 | 0.362 | 0.315 | 0.415 | 0.187 | 0.559 | 0.408 | 0.235 | 0.39 | 0.214 |
| 15+ | 0.348 | 0.277 | 0.36 | 0.36 | 0.36 | 0.35 | 0.336 | 0.389 | 0.353 | 0.402 | 0.375 | 0.404 |  | 0.389 | 0.46 | 0.351 | 0.475 |  | 0.339 | 0.273 |  | 0.378 | 0.26 |

CM LENGTH

| AGE | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 1 | 19.2 | 19.2 | 19.2 | 19.2 | 19.2 | 19.1 | 19.5 | 19.4 | 20.3 | 19.8 | 18.1 | 20.1 | 19.9 | 20 | 20.3 | 20.8 | 19.2 | 19.9 | 20.9 | 20.4 | 19.8 | 20 | 19.1 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 22 | 22 | 22 | 22 | 22 | 21.5 | 21.5 | 21.7 | 22.3 | 22.2 | 21.5 | 22 | 20.8 | 21.6 | 21.6 | 22.6 | 21.7 | 21.7 | 22.4 | 22.9 | 22.9 | 22 | 21.3 |
| 3 | 23.5 | 23.5 | 23.5 | 23.5 | 23.5 | 23.9 | 21.9 | 23.8 | 23.7 | 23.6 | 22.9 | 23.4 | 22.5 | 23.2 | 23.2 | 23.9 | 23 | 23.5 | 23.5 | 23.6 | 24.6 | 23.6 | 23.3 |
| 4 | 24.8 | 24.8 | 24.8 | 24.8 | 24.8 | 24.9 | 23.4 | 25.4 | 24.6 | 25.2 | 24.7 | 24.1 | 23.6 | 24.1 | 24.6 | 25 | 24.5 | 25 | 25.3 | 24.8 | 25.8 | 24.8 | 24.1 |
| 5 | 25.5 | 25.5 | 25.5 | 25.5 | 25.5 | 26 | 26.7 | 26.3 | 26.2 | 26.6 | 25.9 | 25.4 | 24.4 | 25.6 | 25.8 | 25.7 | 25.9 | 25.7 | 27 | 25.4 | 26.6 | 26.4 | 26.7 |
| 6 | 26.4 | 26.4 | 26.4 | 26.4 | 26.4 | 27.6 | 27.5 | 27.4 | 27.3 | 27.5 | 27.7 | 27 | 26.6 | 26.3 | 27.2 | 27.1 | 27.6 | 27 | 27.1 | 27.3 | 28.2 | 26.1 | 27.5 |
| 7 | 27.2 | 27.2 | 27.2 | 27.2 | 27.2 | 28.1 | 28.1 | 28.6 | 28.2 | 28.8 | 29.8 | 28.6 | 27.8 | 28.1 | 28.1 | 28.3 | 27.7 | 27.1 | 28.3 | 27.5 | 30.4 | 27.5 | 27.5 |
| 8 | 29.2 | 29.2 | 29.2 | 29.2 | 29.2 | 28.6 | 28.5 | 29.3 | 29 | 29.2 | 30.4 | 29.8 | 28.1 | 28.8 | 30.6 | 28.4 | 27.8 | 27 | 28.9 | 28 | 31.7 | 30.2 | 28 |
| 9 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 29.9 | 29.8 | 29.4 | 29.9 | 30.4 | 30.8 | 30.8 | 30.1 | 31.2 | 31.1 | 30.2 | 31.9 | 28.6 | 29.2 | 28.8 | 30.5 | 30.5 | 29.1 |
| 10 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 31.2 | 30.2 | 30.3 | 30.9 | 31.4 | 31.8 | 31.5 | 31 | 31.8 | 32.5 | 30 | 32.5 | 28 | 29.5 | 29.2 | 32.5 | 34.7 | 29.5 |
| 11 | 30.6 | 30.6 | 30.6 | 30.6 | 30.6 | 31.5 | 30.7 | 31.4 | 30.7 | 31.9 | 33.8 | 31.2 | 39.5 | 31.6 | 35 | 32.2 | 33.2 | 30.1 | 30 | 30.7 | 31.5 | 35.2 | 31.1 |
| 12 | 32.1 | 32.1 | 32.1 | 32.1 | 32.1 | 33.6 | 32 | 31.6 | 31.9 | 31.7 | 35.6 | 30.8 | 31.5 | 32.2 | 35.3 | 30.8 | 34.6 | 27.5 | 30.4 | 30.6 | 32.3 | 35.5 |  |
| 13 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 31.7 | 32.4 | 32.8 | 31.9 | 34 | 32.1 | 33.4 | 33.9 | 34 | 31.8 | 36.4 |  | 32.1 | 30 | 32.5 | 31.5 |  |
| 14 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 |  | 32.1 | 32.4 | 32.5 | 33 | 34.4 | 32.5 | 34.5 | 32.3 | 34.2 | 33 | 36 | 27.5 | 38.5 | 36 | 30.5 | 36.1 | 30.5 |
| $15+$ | 32.5 | 32.5 | 32.5 | 32.5 | 32.5 | 33.8 | 33.4 | 34.3 | 33.6 | 34.8 | 35.2 | 35.3 |  | 35.1 | 36.1 | 34.5 | 36.9 |  | 34.2 | 32.5 |  | 36.1 | 31.5 |

Table 6.3.3. North Sea Horse Mackerel stock. Mean weight at age (kg) in the catch by quarter and area in 2017.

| Number/1000 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.06 | 0.06 | 0.06 | 0.06 | 0 | 0.06 |
| 2 | 0.086 | 0.086 | 0.086 | 0.086 | 0.058 | 0.081 |
| 3 | 0.114 | 0.114 | 0.114 | 0.114 | 0.087 | 0.108 |
| 4 | 0.13 | 0.13 | 0.13 | 0.13 | 0.112 | 0.127 |
| 5 | 0.172 | 0.172 | 0.172 | 0.172 | 0.151 | 0.168 |
| 6 | 0.19 | 0.19 | 0.19 | 0.19 | 0.159 | 0.184 |
| 7 | 0.175 | 0.175 | 0.175 | 0.175 | 0.163 | 0.173 |
| 8 | 0.188 | 0.188 | 0.188 | 0.188 | 0.178 | 0.186 |
| 9 | 0.222 | 0.222 | 0.222 | 0.222 | 0.225 | 0.223 |
| 10 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 11 | 0.257 | 0.257 | 0.257 | 0.257 | 0.277 | 0.261 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 |
| 15 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 |
| 2Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 |
| 2 | 0.086 | 0.086 | 0.086 | 0.086 | 0.086 | 0.086 |
| 3 | 0.114 | 0.114 | 0.114 | 0.114 | 0.114 | 0.114 |
| 4 | 0.13 | 0.13 | 0.13 | 0.13 | 0.13 | 0.13 |
| 5 | 0.172 | 0.172 | 0.172 | 0.172 | 0.172 | 0.172 |
| 6 | 0.19 | 0.19 | 0.19 | 0.19 | 0.19 | 0.19 |
| 7 | 0.175 | 0.175 | 0.175 | 0.175 | 0.175 | 0.175 |
| 8 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 |
| 9 | 0.222 | 0.222 | 0.222 | 0.222 | 0.222 | 0.222 |
| 10 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 11 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 |


| 15 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0.06 | 0.06 | 0.06 | 0.06 |
| 2 | 0 | 0 | 0.086 | 0.086 | 0.086 | 0.086 |
| 3 | 0 | 0 | 0.114 | 0.114 | 0.114 | 0.114 |
| 4 | 0 | 0 | 0.13 | 0.13 | 0.13 | 0.13 |
| 5 | 0 | 0 | 0.172 | 0.172 | 0.172 | 0.172 |
| 6 | 0 | 0 | 0.19 | 0.19 | 0.19 | 0.19 |
| 7 | 0 | 0 | 0.175 | 0.175 | 0.175 | 0.175 |
| 8 | 0 | 0 | 0.188 | 0.188 | 0.188 | 0.188 |
| 9 | 0 | 0 | 0.222 | 0.222 | 0.222 | 0.222 |
| 10 | 0 | 0 | 0.233 | 0.233 | 0.233 | 0.233 |
| 11 | 0 | 0 | 0.257 | 0.257 | 0.257 | 0.257 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0.214 | 0.214 | 0.214 | 0.214 |
| 15 | 0 | 0 | 0.26 | 0.26 | 0.26 | 0.26 |
| 4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0.06 | 0.06 | 0.059 | 0.074 |
| 2 | 0 | 0 | 0.086 | 0.092 | 0.088 | 0.089 |
| 3 | 0 | 0 | 0.114 | 0.115 | 0.117 | 0.115 |
| 4 | 0 | 0 | 0.13 | 0.141 | 0.148 | 0.14 |
| 5 | 0 | 0 | 0.172 | 0.18 | 0.198 | 0.183 |
| 6 | 0 | 0 | 0.19 | 0 | 0.206 | 0.132 |
| 7 | 0 | 0 | 0.175 | 0 | 0.215 | 0.13 |
| 8 | 0 | 0 | 0.188 | 0 | 0.208 | 0.132 |
| 9 | 0 | 0 | 0.222 | 0 | 0.219 | 0.147 |
| 10 | 0 | 0 | 0.233 | 0 | 0 | 0.078 |
| 11 | 0 | 0 | 0.257 | 0 | 0.234 | 0.164 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0.214 | 0 | 0 | 0.071 |
| 15 | 0 | 0 | 0.26 | 0 | 0 | 0.087 |


| 1-4Q |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.06 | 0.06 | 0.06 | 0.06 | 0.059 | 0.075 |
| 2 | 0.086 | 0.086 | 0.086 | 0.088 | 0.081 | 0.087 |
| 3 | 0.114 | 0.114 | 0.114 | 0.114 | 0.109 | 0.117 |
| 4 | 0.13 | 0.13 | 0.13 | 0.133 | 0.131 | 0.135 |
| 5 | 0.172 | 0.172 | 0.172 | 0.174 | 0.174 | 0.167 |
| 6 | 0.19 | 0.19 | 0.19 | 0.19 | 0.187 | 0.178 |
| 7 | 0.175 | 0.175 | 0.175 | 0.175 | 0.182 | 0.179 |
| 8 | 0.188 | 0.188 | 0.188 | 0.188 | 0.191 | 0.257 |
| 9 | 0.222 | 0.222 | 0.222 | 0.222 | 0.222 | 0.291 |
| 10 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 | 0.359 |
| 11 | 0.257 | 0.257 | 0.257 | 0.257 | 0.255 | 0.369 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.379 |
| 15 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 | 0.37 |

Table 6.3.4. North Sea Horse Mackerel stock. Mean length (cm) at age in the catch by quarter and area in 2017

| Number/1000 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 19.1 | 19.1 | 19.1 | 19.1 | 0 | 19.1 |
| 2 | 21.3 | 21.3 | 21.3 | 21.3 | 18 | 20.7 |
| 3 | 23.4 | 23.4 | 23.4 | 23.4 | 21.5 | 23 |
| 4 | 24.1 | 24.1 | 24.1 | 24.1 | 22.9 | 23.8 |
| 5 | 26.6 | 26.6 | 26.6 | 26.6 | 25.7 | 26.4 |
| 6 | 27.5 | 27.5 | 27.5 | 27.5 | 26.3 | 27.3 |
| 7 | 27.4 | 27.4 | 27.4 | 27.4 | 26.8 | 27.3 |
| 8 | 28 | 28 | 28 | 28 | 27.8 | 28 |
| 9 | 29.1 | 29.1 | 29.1 | 29.1 | 29.6 | 29.2 |
| 10 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 |
| 11 | 31.1 | 31.1 | 31.1 | 31.1 | 32.5 | 31.4 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 |
| 15 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 |
| 2Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 19.1 | 19.1 | 19.1 | 19.1 | 19.1 | 19.1 |
| 2 | 21.3 | 21.3 | 21.3 | 21.3 | 21.3 | 21.3 |
| 3 | 23.4 | 23.4 | 23.4 | 23.4 | 23.4 | 23.4 |
| 4 | 24.1 | 24.1 | 24.1 | 24.1 | 24.1 | 24.1 |
| 5 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 |
| 6 | 27.5 | 27.5 | 27.5 | 27.5 | 27.5 | 27.5 |
| 7 | 27.4 | 27.4 | 27.4 | 27.4 | 27.4 | 27.4 |
| 8 | 28 | 28 | 28 | 28 | 28 | 28 |
| 9 | 29.1 | 29.1 | 29.1 | 29.1 | 29.1 | 29.1 |
| 10 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 |
| 11 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 |


| 15 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 19.1 | 19.1 | 19.1 | 19.1 |
| 2 | 0 | 0 | 21.3 | 21.3 | 21.3 | 21.3 |
| 3 | 0 | 0 | 23.4 | 23.4 | 23.4 | 23.4 |
| 4 | 0 | 0 | 24.1 | 24.1 | 24.1 | 24.1 |
| 5 | 0 | 0 | 26.6 | 26.6 | 26.6 | 26.6 |
| 6 | 0 | 0 | 27.5 | 27.5 | 27.5 | 27.5 |
| 7 | 0 | 0 | 27.4 | 27.4 | 27.4 | 27.4 |
| 8 | 0 | 0 | 28 | 28 | 28 | 28 |
| 9 | 0 | 0 | 29.1 | 29.1 | 29.1 | 29.1 |
| 10 | 0 | 0 | 29.5 | 29.5 | 29.5 | 29.5 |
| 11 | 0 | 0 | 31.1 | 31.1 | 31.1 | 31.1 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 30.5 | 30.5 | 30.5 | 30.5 |
| 15 | 0 | 0 | 31.5 | 31.5 | 31.5 | 31.5 |
| 4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 19.1 | 19.2 | 19.1 | 19.1 |
| 2 | 0 | 0 | 21.3 | 21.9 | 21.6 | 21.6 |
| 3 | 0 | 0 | 23.4 | 23.4 | 23.6 | 23.5 |
| 4 | 0 | 0 | 24.1 | 24.8 | 25.3 | 24.7 |
| 5 | 0 | 0 | 26.6 | 28.5 | 27.5 | 27.5 |
| 6 | 0 | 0 | 27.5 | 0 | 28.1 | 18.5 |
| 7 | 0 | 0 | 27.4 | 0 | 28.7 | 18.7 |
| 8 | 0 | 0 | 28 | 0 | 28.5 | 18.8 |
| 9 | 0 | 0 | 29.1 | 0 | 28.4 | 19.2 |
| 10 | 0 | 0 | 29.5 | 0 | 0 | 9.8 |
| 11 | 0 | 0 | 31.1 | 0 | 29.5 | 20.2 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 30.5 | 0 | 0 | 10.2 |
| 15 | 0 | 0 | 31.5 | 0 | 0 | 10.5 |


| 1-4Q |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 19.1 | 19.1 | 19.1 | 19.1 | 19.1 | 19.1 |
| 2 | 21.3 | 21.3 | 21.3 | 21.5 | 20.7 | 21.3 |
| 3 | 23.4 | 23.4 | 23.4 | 23.4 | 23 | 23.3 |
| 4 | 24.1 | 24.1 | 24.1 | 24.3 | 24.1 | 24.1 |
| 5 | 26.6 | 26.6 | 26.6 | 27.1 | 26.6 | 26.7 |
| 6 | 27.5 | 27.5 | 27.5 | 27.5 | 27.4 | 27.5 |
| 7 | 27.4 | 27.4 | 27.4 | 27.4 | 27.6 | 27.5 |
| 8 | 28 | 28 | 28 | 28 | 28.1 | 28 |
| 9 | 29.1 | 29.1 | 29.1 | 29.1 | 29 | 29.1 |
| 10 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 |
| 11 | 31.1 | 31.1 | 31.1 | 31.1 | 31 | 31.1 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 |
| 15 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 |

Table 6.4.1. North Sea Horse Mackerel. cpue Indices of abundance (individuals/hour) for juvenile $(<20 \mathrm{~cm})$ and exploitable ( $>20 \mathrm{~cm}$ ) substocks, estimated as a combined index for the NS-IBTS Q3 (North Sea only, no 27.7.d included) and the Channel Ground Fish Survey in Q4 (CGFS, 27.7.d). The survey indices are derived from the prediction of a hurdle model fit to data over the period 1992-2017.

|  | Juvenile substock (<20cm) |  |  | Exploitable substock ( $\mathbf{2 0} \mathbf{c m}$ ) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Index | CI_low | CI_high | Index | CI_low | CI_high |
| 1992 | 4865 | 2293 | 9237 | 1498 | 663 | 2915 |
| 1993 | 1917 | 959 | 3415 | 565 | 291 | 974 |
| 1994 | 3288 | 1593 | 6067 | 1322 | 635 | 2368 |
| 1995 | 2232 | 1107 | 4115 | 1621 | 669 | 3401 |
| 1996 | 1178 | 447 | 2480 | 1080 | 482 | 1987 |
| 1997 | 3350 | 1516 | 6253 | 714 | 336 | 1286 |
| 1998 | 858 | 414 | 1573 | 436 | 201 | 806 |
| 1999 | 1475 | 794 | 2433 | 517 | 257 | 905 |
| 2000 | 1139 | 516 | 2333 | 289 | 137 | 570 |
| 2001 | 3431 | 1580 | 7437 | 508 | 245 | 916 |
| 2002 | 2999 | 1515 | 5386 | 501 | 240 | 937 |
| 2003 | 2116 | 1190 | 3499 | 381 | 179 | 726 |
| 2004 | 1064 | 559 | 1844 | 428 | 199 | 754 |
| 2005 | 987 | 530 | 1727 | 759 | 366 | 1370 |
| 2006 | 502 | 271 | 880 | 834 | 422 | 1523 |
| 2007 | 665 | 375 | 1107 | 411 | 197 | 787 |
| 2008 | 394 | 224 | 695 | 209 | 101 | 458 |
| 2009 | 758 | 416 | 1265 | 104 | 47 | 212 |
| 2010 | 1611 | 863 | 2981 | 234 | 106 | 459 |
| 2011 | 569 | 317 | 1091 | 282 | 136 | 583 |
| 2012 | 354 | 189 | 705 | 185 | 93 | 437 |
| 2013 | 1062 | 572 | 1882 | 146 | 64 | 335 |
| 2014 | 1609 | 909 | 2747 | 430 | 193 | 876 |
| 2015 | 2257 | 1220 | 4527 | 580 | 261 | 1146 |
| 2016 | 1752 | 959 | 2976 | 803 | 376 | 1557 |
| 2017 | 973 | 505 | 1714 | 131 | 54 | 282 |

### 6.9 Figures

TAC and catch uptake. Year 2017


Figure 6.2.1. North Sea horse mackerel. Utilisation of quota by country.


Figure 6.2.2 North Sea Horse Mackerel. North Sea horse mackerel. Catch by ICES Division for 20002017.


Figure 6.2.3.- North Sea Horse Mackerel. Total catch (in tonnes) by area, quarter, catch category and country. BMS landing refers to landings below minimum legal size.

North Sea Stock: Catch by division


Figure 6.3.1.- North Sea Horse Mackerel. Proportion of NSHM total catch per year and station that have been sampled.


Figure 6.3.2.- North Sea horse mackerel age distribution in the catch for 1995-2017. The area of bubbles is proportional to the catch number. Note that age 15 is a plus group.


NSHM: catch at age ( N ; observed) out of 27.7.d


Figure 6.3.3. North Sea horse mackerel. Bubbleplots of age distribution in the catch by area for 19982017 for area 7.d (upper panel) and out of 7.d (bottom panel). The area of bubbles is proportional to the total catch number for the stock. Note that age 15 is a plus group.


Figure 6.3.4. North Sea horse mackerel. Mean weight at age in commercial catches over the period 2000-2017

## Mean length at age (cm)



Figure 6.3.5. North Sea horse mackerel. Mean length at age in commercial catches over the period 2000-2017.


Figure 6.4.1. North Sea Horse Mackerel. Catch curves for the 1994 to 2007 cohorts, ages from 3 to 14. Values plotted are the log(catch) values for each cohort in each year. The negative slope of these curves estimates total mortality $(\mathrm{Z})$ in the cohort.

Total mortality by cohort


Figure 6.4.2. North Sea Horse Mackerel. Total mortality by cohort (Z) estimated from the negative gradients of the 1992-2006 cohort catch curves (Figure 6.4.1).


Figure 6.4.3.- North Sea horse mackerel. ICES rectangles selected in 2013 and currently used by the working group.


Figure 6.4.4.- North Sea horse mackerel. Size distribution of North Sea horse mackerel catches during the inter-calibration exercise conducted in 2014 between the RV Gwen Drez (red bars) and Thalassa (blue bars).


Figure 6.4.5. North Sea horse mackerel. cpue by depth for the CGFS survey from 1992 to 2017.


Figure 6.4.6. North Sea Horse Mackerel. Combined cpue survey index (indiv/hour) derived from the hurdle model fit to the IBTS survey in the North Sea (4.bc) and the CGFS survey in the English channel. Top: Juvenile substock ( $<20 \mathrm{~cm}$ ); Bottom: exploitable substock ( $>20 \mathrm{~cm}$ ). The abundance index values are presented as number of individuals per hours. The confidence interval is determined by bootstrap resampling of Pearson residuals with 1000 iterations.


Figure 6.4.7.- North Sea horse mackerel. Proportion of hauls with zero catch for the exploitable ( $\mathbf{~} 20 \mathrm{~cm}$ ) and juvenile ( $<20 \mathrm{~cm}$ ) substocks in the NS-IBTS (pink dotted lines) and the CGFS (blue dotted lines).

$>20 \mathrm{~cm}$ substock


Figure 6.4.8. North Sea Horse Mackerel. Mean CPUE survey index (indiv/hour) obtained from the hurdle model fit to the IBTS survey in the North Sea (4.bc) and the CGFS survey in the English channel. Top: Juvenile substock ( $<20 \mathrm{~cm}$ ); Bottom: exploitable substock ( $>20 \mathrm{~cm}$ ). The abundance index values are presented as number of individuals per hours.


Figure 6.4.9. North Sea horse mackerel. Relative occurrence by length for the period 2012-2017 in the CGFS.


Figure 6.4.10. North Sea horse mackerel. Relative occurrence by length for the period 2012-2017 in the IBTS.


Figure 6.4.11.- Length distribution, as well as the estimated parameters Lc, Lmean, $L f=m$ for the Dutch fleet in 2016 and 2017.

## 6 North Sea Horse Mackerel: Divisions 27.4.a (Q1 and Q2), 27.3.a (excluding Western Skagerrak Q3 and Q4), 27.4.b, 27.4.c and

 27.7.d
### 6.1 ICES Advice Applicable to 2018

In 2012 the North Sea horse mackerel (NSHM) was classified as a category 5 stock, based on the ICES approach to data-limited stocks (DLS). Since then, a progressive reduction of TAC was advised by ICES, from 25500 tonnes in 2013-2014 to 15200 tonnes in 2015-2016. This reduction in the advised catch was supported by the analysis of information from the North Sea International Bottom Trawl Survey (NS-IBTS) traditionally used in the assessment, but also new information from the Channel Ground Fish Survey (CGFS) since 2014. Despite the considerable increase showed by the CGFS in 2015 survey, due to the high uncertainty in the two survey indices, catch in 2017 was advised to continue at 15200 tonnes. However, new information indicated a $16.7 \%$ discards of NSHM in 2015 in non-directed fisheries. This new information is taken into account in the catch advice for 2017. The advice landings were 15200 tonnes and the advice total catch was 18247 tonnes.
In 2017 this stock was benchmarked and the NS-IBTS and CGFS survey indices where modelled together. The resulting joint index was considered a proper indication of trend in abundance over time and the NSHM stock was upgraded to category 3 . The joint survey index showed in 2016 a continuation of the increasing trend started in 2013. The application of the HCR 3.1 (ICES, 2012, comparison of the two latest index values with the three preceding values multiplied by the recent advised catch) resulted in an increase higher than $20 \%$ in the catch advice for 2018 in comparison to advice for 2017. Accordingly the uncertainty cap was applied. In addition, Length Based DLS methods indicated that the F was in 2016 slightly above the Fmsy proxy, and stock size relative to reference points was unknown. Therefore, the precautionary buffer was also applied to the advice, since it hadn't been applied since 2014. This resulted in a catch advice for 2018 and 2019 no more than 17517 tonnes. Considering the $13.35 \%$ average discards, were advice not being higher than 15179 tonnes.

### 6.2 Fishery of North Sea horse mackerel stock

Based on historical catches taken by the Danish industrial fleet for reduction into fishmeal and fish oil in the 1970s and 1980s, approximately $48 \%$ of the EU North Sea horse mackerel TAC was taken by Denmark. Catches were taken in the fourth quarter mainly in Divisions 27.4.b and 27.7.d. The 1990s saw a drop in the value of industrial fish, limited fishing opportunities and steep increases in fuel costs that affected the Danish quota uptake. In 2001, individual quota scheme for a number of species was introduced in Denmark, but not for North Sea horse mackerel. This lead to a rapid restructuring and lower capacity of the Danish fleet, which in combination with the above mentioned factors led to a decrease of the Danish North Sea horse mackerel catches.

Since the 1990's, a larger portion of catches has been taken in a directed horse mackerel fishery for human consumption by the Dutch freezer-trawler fleet. This is possible because Denmark has traded parts of its quota with the Netherlands for other species. However due to the structure of the Danish quota management setup only a limited amount of quota can be made available for swaps with other countries. These practical implications of the management scheme largely explain the consistent underutilisation of the TAC over the period 2010-2013 (approximately $50 \%$ ). However, following the sharp reduction in TAC in 2014 uptake increased significantly to above $80 \%$ in 2015
and $100 \%$ in 2016 (see Figure 6.2.1), although an important part of these catches were discards ( $16.7 \%$ and $10 \%$ respectively). In 2017 the $80 \%$ of the TAC was used, with an 8.3\% discards.

Catches taken in Divisions 27.3.a and 27.4.a during the two first quarters and all year in Divisions 27.4.b, 4.c and 27.7.d are regarded North Sea horse mackerel (Section 5, Table 5.4.1). The catches were relatively low during the period 1982-1997 with an average of 18000 tonnes, but increased between 1998 (30500 tonnes) and 2000 (45130 tonnes). From 2000 to 2010, the catches varied between 24149 and 45883 tonnes. Since 2014 a steep decline in catches is observed, both due to the reduction in the TAC since 2014 but also the underutilization of the quota. In 2017 the catch was 14579 tonnes, with an $82 \%$ of total catch being caught in area 27.7.d.

Over the period 1985-2001 most catches were taken in the area 27.4.b (Figure 6.2.3). However, since 2002 the proportion of catches from area 27.7.d increased steadily until 2013, when the $92 \%$ of total catches were fished in this area (Figure 6.2.2). Germany, UK and Netherlands accounted for most of the landings, that were taken in quarter 1, but especially in quarter 4 (Figure 6.2.3). Most of the discards were reported in area 7d, more importantly during quarters $3^{\text {rd }}$ and $4^{\text {th }}$, by the French bottom-trawl fleet. Discarding in the target pelagic fisheries is considered negligible. New information in 2015 from bottom-trawl fisheries not directed at horse mackerel indicated an overall discard rate of $16.7 \%$ for the stock as a whole, while in 2016 this rate is $10 \%$. Complete discard information for earlier years has not been submitted to ICES. However, information from national discard reports for the non-directed bottom-trawl fisheries indicates a similar level of discarding in earlier years.

### 6.3 Biological Data

### 6.3.1 Catch in Numbers at Age

In 2017, as already occurred in 2016, there has been a marked reduction in the coverage of biological sampling. Only the $38 \%$ of landings was sampled, in comparison to 2013 and 2014 when $71 \%$ and $63 \%$ were sampled respectively (section 5 figure 5.9.1). In addition, this low coverage was carried out mostly by the Dutch fleet in quarter Q4 and divisions 27.7.d and 27.4.c. Despite most landing catch was taken from this area and quarter (81.9\% of landings in division 27.7.d and $75 \%$ in quarter Q4, Figure 6.3.1) still part of landings were fished in other areas and quarters. In order to avoid a biased perception of the age distribution of catches over the year and areas, this partial and uneven sampling effort should be avoided in future years.

Annual catch numbers at age by area for year 2017 are shown in table 6.3.1. Due to the low level of sampling effort out of area 27.7.d., there is not enough information to represent age distribution in those areas, and hence, the one observed in 27.7.d is taken as the basis to separate catch by age. Catch-at-age for the whole period 1995-2017 are given in, table 6.3.2 and in Figure 6.3.2. These data show that since 2005 the age distribution of catches has experienced a reduction, with a decrease in the range of ages of importance in total catches. In parallel to the rejuvenation of catches, the comparison of catch-at-age data after 1998 by area (Figure 6.3.3) shows that since 2010 commercial catches have increased in area 27.7.d in comparison to the areas 27.3.a and $4 \mathrm{a}, \mathrm{b}$ and c where the opposite pattern was found. Since 2015, commercial catches are focused mostly on cohort 2014 that was the main component of catches both in and out of the area 27.7.d at age 1 in year 2015, age 2 in 2016 and age 3 in 2017. Ages 1 and 2 appear with moderate importance in the total catch.

Although 2015 cohort seems to be clear in the catch-at-age distribution, in general cohort structure is not clearly detectible in the data. This may partly be due to the shifts in distribution of the fishery. In addition, it may partly be due to age reading difficulties, which are a known to be encountered (e.g. Bolle et al., 2011). Most clearly detectable is the relatively large 2001 year class, although it is not clearly present in the catch in all years. There are indications that environmental circumstance may be an important factor (possibly stronger than stock size) contributing to spawning success in horse mackerel. This is for example illustrated by the largest year classes (1982 and 2001) observed in the Western stock which incidentally were produced at the lowest observed stock sizes. Since 2001 is considered to have been a relatively strong year class in the Western stock as well, it is plausible that circumstances in the North Sea were similar to those in Western areas and also allowed for relatively high spawning success in the North Sea.

Lastly, potential mixing of fish from the Western and North Sea stock in area 27.7.d and 27.7.e in winter may also confuse the cohort signals. For example, the large recruitment in the Western stock may have led to more of these fish being located in the North Sea stock area as age 1 fish in 2002. With the intention of clarifying the mixing among the North Sea and the Western horse mackerel stocks, and how this may affect to the age distribution of catches. In 2015 was conducted by IMARES and the Pelagic Freezertrawler Association (PFA). The results of this project were not conclusive because of the low sample size and some technical glitches in the sequencing. The chemical analysis generated some new insights but also some more questions on the variability that could be expected between years and seasons. Currently more genetic samples are being taking by PFA in different areas of the North East Atlantic. In addition, catch sampling carried out by several pelagic fishery companies is being explored to give information on the separation between North Sea and Western horse mackerel. Until the mixing of both stocks is clarified and catch-at-age data can be clearly segregated the development of analytical assessment will be limited.

### 6.3.2 Mean weight at age and mean length at age

The mean weight and mean length-at-age in the commercial catches of 2017 are presented in tables 6.3.3 and 6.3.4 respectively by quarter and area. As explained for the distribution of catch-at-age by area, due to the biased sampling coverage in 2017 for several ages mean weight and length in quarters Q2-Q3 in areas 27.3.a, 24.7.a-b-c are assumed the same than in quarter Q1-Q4 in area 27.7.d.

The mean annual weight and length over the period 2000-2016 are presented in table 6.3.2 and figures 6.3.4 and 6.3.5 respectively. Despite there are no strong differences over this period, since 2006 there seems to be a slight but steady increase in both weight and length until 2015, when a declining pattern is observed. It may be hypothesized that this is due to density-dependent effects, due to the relatively successful recruitment of 2015.

### 6.3.3 Maturity-at-age

Peak spawning in the North Sea falls in May and June (Macer, 1974), and spawning occurs in the coastal regions of the southern North Sea along the coasts of Belgium, the Netherlands, Germany, and Denmark.

There is no information available about the maturity-at-age of the North Sea Horse mackerel stock.

### 6.3.4 Natural mortality

There is no specific information available about natural mortality of this stock.

### 6.4 Data Exploration

### 6.4.1 Catch curves

The log-catch numbers were plotted by cohort to estimate the negative gradient of the slope and get an estimate of total mortality (Z). Fully selected ages 3 to 14 from the 1992 - 2016 period provide complete data for the 1992 to 2006 cohorts (Figure 6.4.1). The estimated negative gradients by cohort (Figure 6.4.2) indicate an increasing trend in total mortality for the period examined, with a marked increment in the cohorts 2005 and 2006. However, due to the low quality of the signals for some cohorts these Z estimates has to be considered with caution.

An analysis of the catch number at age data carried out in 2011 showed that only the $1 \mathrm{vs} .2,2 \mathrm{vs} .3,7 \mathrm{vs} .8$ and 8 vs .9 age groups were positively and significantly correlated in the catch. This analysis was not updated this year but these results suggest limitations in the catch-at-age data.

### 6.4.2 Assessment models and alternative methods to estimate the biomass

In 2002 Ruckert et al. estimated the North Sea horse mackerel biomass based on a ratio estimate that related cpue data from the IBTS to cpue data of whiting (Merlangius merlangus). The applied method assumes that length specific catchability of whiting and horse mackerel are the same for the IBTS gear. Subsequently, they use the total biomass of whiting derived from an analytical stock assessment (MSVPA) to estimate the relationship between cpue and biomass.

At the 2014 WGWIDE some exploratory model fits were attempted with the JAXass model, using the data available. The JAXass (JAX assessment) model is a simple statistical catch-at-age model fitted to an age-aggregated index of ( $2+$ ) biomass, total catch data and proportions at age from the catch. It is based on Per Sparre's "separable VPA" model, an ad hoc method tested for the first time at WGWIDE in 2003, and later 2004. A new analysis using this model was also done in 2007 using an IBTS index. In 2014 the model has been coded in ADMB (Fournier et al., 2012) and updated with an improved objective function (dnorm), extra years of data and new methods for calculating the index (see above).

Difficulties in fitting an assessment model for this stock include:

- Unclear stock boundaries
- Difficulty aging horse mackerel
- Lack of strong cohort signals in CAA data.
- Scientific index derived from a survey not specifically designed for horse mackerel and not covering one of the main fishing grounds for the stock (VIId)

Catches taken in area 27.7. d are close to the management boundary between the (larger) western horse mackerel stock and the NS horse mackerel stock. It is quite possible that given changes in oceanographic condition, or changes in abundance of either of the two stocks, that some proportion of the catches taken in area 27.7.d actually originated from the western horse mackerel stock. Nevertheless, all assessment models
used in the MSE assume that $100 \%$ of fish caught in area 27.7.d belong to the North Sea horse mackerel stock. This is in agreement with stock and management definitions.

### 6.4.3 Survey data

### 6.4.3.1 Egg Surveys

No egg surveys for horse mackerel have been carried out in the North Sea since 1991. Such surveys were carried out during the period 1988-1991. SSB estimates are available historically. However, they were calculated assuming horse mackerel to be a determinate spawner. Horse mackerel is now considered an indeterminate spawner. Therefore egg abundance could only be considered a relative index of SSB. The mackerel egg surveys in the North Sea do not cover the spawning area of horse mackerel.

### 6.4.3.2 IBTS Survey Data

Many pelagic species are frequently found close to the bottom during daytime (which is when the IBTS survey operates) and migrate upwards predominantly during the night they are susceptible to semi-pelagic fishing gear and to bottom trawls (Barange et al. 1998). Eaton et al. (1983) argued that horse mackerel of 2 years and older are predominantly demersal in habit. Therefore, in the absence of a targeted survey for this stock, the IBTS is considered a reasonable alternative. IBTS data are also used in the assessment of the southern horse mackerel stock.

IBTS data from quarter Q3 were obtained from DATRAS and analysed. Based on a comparison of IBTS data from all 4 quarters in the period 1991-1996, Ruckert et al. (2002) showed that horse mackerel catches in the IBTS were most abundant in the third quarter of the year. In 2013 WGWIDE considered that using an 'exploitable biomass index' estimated with the abundance by haul of individuals larger than 20 cm is the most appropriate for the purpose of interpreting trend in the stock.

To create indices, a subset of ICES rectangles was selected. Rectangles that were not covered by the survey more than once during the period 1991-2012 were excluded from the index area. In 2012, WGWIDE expressed concern that the previously selected index area did not sufficiently cover the distribution area of the stock, especially in years that the stock would be relatively more abundant and spread out more. Ruckert et al. (2002) also identified a larger distribution area of the North Sea stock. Based on the above, 61 rectangles were identified to be included in the index area as shown in Figure 6.4.3.

### 6.4.3.3 The French Channel Groundfish Survey (CGFS) in Q4

In order to improve data basis for the North Sea horse mackerel assessment, alternative survey indices have been explored. Previous indices used had only cover the North Sea distribution of the stock, while the majority of catches in recent years have come from the eastern English Channel (27.7.d). We evaluated the potential contribution of the French Channel Ground Fish Survey in 27.7.d (CGFS) in Quarter 4. The CGFS is carried out since 1990 and has frequent captures of horse mackerel. Though this survey is conducted in a different quarter to the North Sea IBTS, the observed seasonal migration patterns of horse mackerel indicate that fish move into the channel following quarter Q3, so the timing is considered appropriate.

In 2015, the RV "Gwen Drez" was replaced by the RV "Thalassa" to carry out the CGFS survey. In 2014 an inter-calibration process was conducted to quantify the differences
in catchability for a large number of species. ICES reviewed this inter-calibration exercise and found a number of drawbacks that may undermine the reliability of the estimated conversion factors. The main concerns were:

- The analyses were limited in the number of tows. Considering that a number of these tows could be zeros for one of the two vessels and possibly resulting in highly uncertain estimates.
- Lack of length-specific correction factor.
- At a standardized depth of 50 m and above, wing spread estimates for the R/V Thalassa as measured by the MARPORT sensor were deemed erroneous, which may question the validity of estimated area swept by the net on the R/V Thalassa and the effect it may have on correction factors for species caught at depth at 50 m and greater.
- A number of tow locations including areas outside 27.7.d were excluded. Changing the depth range of a survey can add serious bias in the calibration and the current approach seems to be ignoring this issue.
- Correction coefficients were not measured without error.

However, these limitations were considered by WGWIDE to be of minor importance for the North Sea horse mackerel since:

- Despite being still a low sample size the North Sea horse mackerel was present in all the 32 paired hauls.
- There are no important differences in size distribution (Figure 6.4.4).
- The analysis with and without the areas excluded in the new sampling design did not show important differences (ICES, 2017).
- cpue or North Sea horse mackerel for hauls deeper than 50 m was relatively low (Figure 6.4.5), and it is expected than the potential problems in determining the conversion factor bellow that depth range would have a relatively minor impact in the estimated abundance.

For these reasons it was finally decided to continue using the CGFS survey, standardizing the time-series of abundance for the period 1990-2015 with the estimated conversion factor 10.363.

### 6.4.3.4 Norht Sea horse mackerel benchmark exercise

In January 2017, a benchmark process was conducted for NSHM (ICES, 2017). Based on capacity to model the overdispersion and the high proportion of zero values in the survey catch data the hurdle models was concluded the best option of all the model alternatives tested. The log-likelihood ratio test, the AIC and the evidence ratio statistic supported that the model that best represented the data was a hurdle model with Year and Survey as explanatory factors (including the interaction term) in the count model (GLM-negative binomial), and Year and Survey (without the interaction) in the zero model (GLM-binomial).

The probability of having a cpue zero was modelled by a logistic regression with a GLM-binomial distribution model:

$$
\operatorname{logit}\left(\pi_{i}\right)=\text { Intercept }_{z e r o}+\text { Year }_{i, z e r o}+\text { Survey }_{i, z e r o}
$$

Where $\pi_{i}$ is the mean probability of having a cpue zero as a function Year and Survey.

The expected cpue of North Sea horse mackerel, conditional to not having a zero in hurdle models (not having a false zero in zero inflated models) was modelled with a GLM-negative binomial distribution model:

$$
\log \left(C P U E_{i}\right)=\text { Intercept }_{\text {count }}+\text { Year }_{i, \text { count }} x \text { Survey }_{i, \text { count }}
$$

This model was used to synthesise the information from both the GCFS and IBTS and predict the average annual cpue index per haul as an indicator of trends in stock abundance both for the juvenile $(<20 \mathrm{~cm})$ and exploitable $(>20 \mathrm{~cm})$ substocks. The contribution of the two surveys to the combined index is weighted taken into consideration their respective surface coverage as well as the mean wing spread. This index model allowed upgrading the NSHM to a category 3 stock within the ICES classification.

### 6.4.4 Summary of index trends

The survey index for both the small and exploitable substocks experienced a marked decline in the early-mid 2000s (Figure 6.4.6; table 6.4.1). This reduction was due in part to the decline of the average abundance by haul over time, but also to the increase of hauls with zero catch of horse mackerel, from $26 \%$ in 1998 to the highest observed value of $72 \%$ in 2013 (Figure 6.4.7). Since 2014 a slight decrease in zero hauls was observed in juveniles group (smaller than 20 cm ). Since 2013, in addition to the decline of zero hauls, the mean cpue in the non-zero hauls has increased. After an increase in 2016, the abundance survey index for the exploitable substock has shown a marked decline in 2017. This pattern has been mostly due to the decline of the survey index estimated for the CGFS in comparison to the value in 2016 (Figure 6.4.8). The survey index of the juvenile substock, that also showed an increasing pattern since 2013, seems to be stabilized since 2014 in the CGFS, but in the IBTS is in 2017 at the lowest level since 1992. Due to this compensation by the CGFS, the abundance index for juveniles, show a less steep decline than the exploitable substock index. The size distribution in both the CGFS and the IBTS suggest the entrance of a moderate new cohort in 2017 (between 47 cm in the IBTS and $7-11 \mathrm{~cm}$ in the CGFS) age 0 (Figures 6.4.9 and 6.4.10).

However, despite the index of abundance of individuals smaller than 20 cm could be considered a recruitment index, it has to be considered with caution. Preliminary examinations of how the juvenile $(0-19 \mathrm{~cm})$ indices relate to subsequent exploitable abundance $(20+\mathrm{cm})$ do not indicate strong linkages. The very high juvenile indices in the early 2000s in the IBTS were not subsequently picked up in the exploitable component. Hence while increases in the juvenile indices are encouraging, whether these lead to increases in the exploitable component of the stock need to be confirmed in the future with observations in the $20+\mathrm{cm}$ indices.

### 6.4.5 Data Limited Stock methods and MSY proxy reference points

As part of the ICES approach to provide advice within the MSY framework for stocks of category 3 and 4, different Data Limited methods to estimate MSY proxy reference points for the North Sea horse mackerel were explored. This analysis and results were presented in a separate working document (Pérez-Rodríguez, 2017). After exploring the compliance with each method assumptions and assessing the data availability the group decided that the Length Based Indicators would be the only DLS method to be applied to the NSHM.

Despite this length based method will have to be applied in the future to a longer timeseries of catch length frequencies, only length data have been collected for 2016 and 2017. The estimates of F/Fmsy proxy indicate that fishing mortality seems to be, both in

2016 and 2017, slightly above FMSY for the North Sea horse mackerel, with F/FMSY=1.082 and 1.073 respectively (Figure 6.4.11).

### 6.4.6 Ongoing work

To improve the knowledge base for North Sea horse mackerel about the degree of connection and migrations in between the North Sea and the Western Stock, catch sampling carried out by several pelagic fishery companies is being explored to give information on the separation between North Sea and Western horse mackerel. To improve the abundance indicators the potential application of a commercial fishery search time index will be explored. Horse mackerel is fished while it is very close to the bottom in relatively dispersed, small schools. The fishery is mostly executed using long hauls and there may be extensive search time involved. Handled in an appropriate statistical framework, taking into account the nature of the fishery and other factors such as seasonality and alternative fishing opportunities, the search time and catch rates could provide for an indication of changes in stock size over time. Catch rates in areas 27.7.e, 27.7.d and southern North Sea will be analysed from skippers' private logbooks.

### 6.5 Basis for 2019 Advice. ICES DLS approach.

Stock advice for NSHM is biannual. In 2017 the advice for years 2018 and 2019 was provided. The joint abundance survey index indicated a continuation in the increasing trend observed since 2013. This increase was due mostly to the increment observed in the CGFS survey index. Despite that, as mentioned in the previous section, the joint survey index for 2017 has shown a sudden change and steep decline due to the drop of the CGFS survey index, WGWIDE decided to continue with the current advice of 17517 tonnes for 2019.

The fisheries in the area have largely been focused on the smaller fish in 2015, 2016, 2017 and it is expected that this will continue in 2018 and 2019. With this pattern of exploitation, mostly immature individuals are caught which might hinder the recovery of the stock by removing an important portion of the recent year classes before they enter the spawning stock.

### 6.6 Management considerations

In the past, Division 27.7.d was included in the management area for Western horse mackerel together with Divisions 27.2.a, 27.7.a-c, 27.7.e-k, 27.8.a, 27.8.b, 27.8.d, 27.8.e, Subarea 6, EU and international waters of Division 5.b, and international waters of Subareas 12 and 14. ICES considers Division 27.7.d to be part of the North Sea horse mackerel distribution area. Since 2010, the TAC for the North Sea area has included Divisions 27.4.bc and 27.7.d. Considering that a majority of the catches are taken in Division 27.7.d, the total of North Sea horse mackerel catches are effectively constrained by the TAC since the realignment of the management areas in 2010.

Catches in Divisions 27.3.a (Western Skagerrak) and 27.4.a in quarters 3 and 4 are considered to be from the Western horse mackerel stock, while catches in quarters 1 and 2 are considered to be from the North Sea horse mackerel stock. Catches in area 27.4.a and 27.3.a are variable. In recent years only Norway has had significant catches in this area, but these are only taken in some years.

### 6.7 References

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### 6.8 Tables

Table 6.3.1. North Sea Horse Mackerel stock. Catch in numbers (1000) by quarter and area in 2017.

| Number/1000 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.99 | 44.35 | 96.06 | 10.51 | 0 | 151.91 |
| 2 | 0.9 | 40.45 | 87.61 | 9.59 | 1579.74 | 1718.29 |
| 3 | 2.73 | 122.35 | 265.02 | 29 | 6243.76 | 6662.86 |
| 4 | 0.4 | 17.94 | 38.85 | 4.25 | 4502.28 | 4563.72 |
| 5 | 0.25 | 11.08 | 24 | 2.63 | 3308.95 | 3346.91 |
| 6 | 0.15 | 6.87 | 14.89 | 1.63 | 1229.79 | 1253.33 |
| 7 | 0.11 | 4.99 | 10.8 | 1.18 | 2074.86 | 2091.93 |
| 8 | 0.1 | 4.45 | 9.63 | 1.05 | 1604.79 | 1620.02 |
| 9 | 0.04 | 1.75 | 3.78 | 0.41 | 558.59 | 564.58 |
| 10 | 0.01 | 0.49 | 1.06 | 0.12 | 261.55 | 263.23 |
| 11 | 0.02 | 1.02 | 2.2 | 0.24 | 297.05 | 300.53 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.01 | 0.49 | 1.06 | 0.12 | 261.55 | 263.23 |
| 15 | 0.01 | 0.49 | 1.06 | 0.12 | 261.55 | 263.23 |
| Sum | 5.72 | 256.7 | 556.04 | 60.84 | 22184.47 | 23063.76 |
| 2Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 16.65 | 24.39 | 17.35 | 411.9 | 828.12 | 1298.41 |
| 2 | 15.18 | 22.24 | 15.83 | 375.67 | 755.27 | 1184.19 |
| 3 | 45.93 | 67.28 | 47.87 | 1136.4 | 2284.7 | 3582.19 |
| 4 | 6.73 | 9.86 | 7.02 | 166.61 | 334.96 | 525.18 |
| 5 | 4.16 | 6.09 | 4.34 | 102.93 | 206.94 | 324.46 |
| 6 | 2.58 | 3.78 | 2.69 | 63.83 | 128.32 | 201.19 |
| 7 | 1.87 | 2.74 | 1.95 | 46.31 | 93.1 | 145.97 |
| 8 | 1.67 | 2.44 | 1.74 | 41.29 | 83.02 | 130.17 |
| 9 | 0.66 | 0.96 | 0.68 | 16.22 | 32.61 | 51.13 |
| 10 | 0.18 | 0.27 | 0.19 | 4.55 | 9.15 | 14.35 |
| 11 | 0.38 | 0.56 | 0.4 | 9.45 | 18.99 | 29.77 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |


| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14 | 0.18 | 0.27 | 0.19 | 4.55 | 9.15 | 14.35 |
| 15 | 0.18 | 0.27 | 0.19 | 4.55 | 9.15 | 14.35 |
| Sum | 96.37 | 141.15 | 100.44 | 2384.26 | 4793.5 | 7515.73 |


| 3Q |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 59.57 | 59.57 |
| 1 | 0 | 0 | 70.95 | 111.67 | 1329.97 | 1512.59 |
| 2 | 0 | 0 | 64.71 | 101.85 | 1212.97 | 1379.53 |
| 3 | 0 | 0 | 195.75 | 308.08 | 3669.25 | 4173.08 |
| 4 | 0 | 0 | 28.7 | 45.17 | 537.94 | 611.81 |
| 5 | 0 | 0 | 17.73 | 27.9 | 332.34 | 377.98 |
| 6 | 0 | 0 | 10.99 | 17.3 | 206.08 | 234.38 |
| 7 | 0 | 0 | 7.98 | 12.55 | 149.52 | 170.05 |
| 8 | 0 | 0 | 7.11 | 11.19 | 133.33 | 151.64 |
| 9 | 0 | 0 | 2.79 | 4.4 | 52.38 | 59.57 |
| 10 | 0 | 0 | 0.78 | 1.23 | 14.7 | 16.72 |
| 11 | 0 | 0 | 1.63 | 2.56 | 30.5 | 34.68 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0.78 | 1.23 | 14.7 | 16.72 |
| 15 | 0 | 0 | 0.78 | 1.23 | 14.7 | 16.72 |
| Sum | 0 | 0 | 410.69 | 646.39 | 7757.96 | 8815.04 |
| 4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 391.05 | 391.05 |
| 1 | 0 | 0 | 6.18 | 1874.01 | 15901.14 | 17781.33 |
| 2 | 0 | 0 | 5.64 | 4425.6 | 12187.97 | 16619.21 |
| 3 | 0 | 0 | 17.06 | 11524.87 | 36654.11 | 48196.04 |
| 4 | 0 | 0 | 2.5 | 1611.53 | 2872.44 | 4486.47 |
| 5 | 0 | 0 | 1.55 | 250.28 | 1730.29 | 1982.11 |
| 6 | 0 | 0 | 0.96 | 0 | 1722.52 | 1723.48 |
| 7 | 0 | 0 | 0.7 | 0 | 413.71 | 414.4 |
| 8 | 0 | 0 | 0.62 | 0 | 542.37 | 542.99 |
| 9 | 0 | 0 | 0.24 | 0 | 263.86 | 264.1 |
| 10 | 0 | 0 | 0.07 | 0 | 0 | 0.07 |
| 11 | 0 | 0 | 0.14 | 0 | 173.56 | 173.7 |


| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0.07 | 0 | 0 | 0.07 |
| 15 | 0 | 0 | 0.07 | 0 | 0 | 0.07 |
|  | 0 | 0 | 786.34 | 999.45 | 80108.12 | 81893.91 |
| 1-4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 450.61 | 450.61 |
| 1 | 17.64 | 68.73 | 190.55 | 2408.1 | 18059.23 | 20744.25 |
| 2 | 16.09 | 62.69 | 173.79 | 4912.7 | 15735.95 | 20901.21 |
| 3 | 48.66 | 189.63 | 525.7 | 12998.35 | 48851.82 | 62614.16 |
| 4 | 7.13 | 27.8 | 77.07 | 1827.55 | 8247.61 | 10187.17 |
| 5 | 4.41 | 17.18 | 47.62 | 383.74 | 5578.52 | 6031.46 |
| 6 | 2.73 | 10.65 | 29.53 | 82.76 | 3286.72 | 3412.39 |
| 7 | 1.98 | 7.73 | 21.42 | 60.04 | 2731.19 | 2822.36 |
| 8 | 1.77 | 6.89 | 19.1 | 53.54 | 2363.51 | 2444.81 |
| 9 | 0.69 | 2.71 | 7.5 | 21.03 | 907.44 | 939.38 |
| 10 | 0.19 | 0.76 | 2.11 | 5.9 | 285.41 | 294.37 |
| 11 | 0.4 | 1.58 | 4.37 | 12.25 | 520.09 | 538.69 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.19 | 0.76 | 2.11 | 5.9 | 285.41 | 294.37 |
| 15 | 0.19 | 0.76 | 2.11 | 5.9 | 285.41 | 294.37 |
| Sum | 102.09 | 397.85 | 1102.97 | 22777.77 | 107588.92 | 131969.6 |

Table 6.3.2. Numbers at age (millions), weight at age (kg) and length at age (cm) for the North Sea horse mackerel 1995-2017 in the commercial fleet catches.

| MILLIO |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 201 | 2017 |
| 1 | 1.8 | 4.6 | 12.6 | 2.3 | 12.4 | 70.2 | 12.8 | 60.4 | 13.8 | 15.7 | 52.4 | 5 | 3.4 | 1.7 | 34.1 | 3.3 | 8.1 | 9.5 | 7.6 | 15.4 | 49.7 | 3.6 | 20.7 |
| 2 | 3.1 | 13.8 | 27.2 | 22.1 | 31.5 | 78 | 36.4 | 16.8 | 56.2 | 17.5 | 29.8 | 23.7 | 15.5 | 8.8 | 13.9 | 22.5 | 23.3 | 24.3 | 10 | 15.3 | 23.8 | 65.2 | 20.9 |
| 3 | 7.2 | 11 | 14.1 | 36.7 | 23.1 | 28.4 | 174.3 | 19.3 | 23.4 | 34.4 | 27.8 | 61.5 | 22.8 | 36.1 | 28.4 | 10.7 | 76.5 | 20.4 | 21.3 | 8.7 | 10.1 | 15.9 | 62.6 |
| 4 | 10.3 | 11.9 | 14.9 | 38.8 | 17.6 | 21.4 | 87.8 | 11.9 | 33.2 | 14.5 | 12.6 | 40.9 | 82.6 | 16.7 | 22.1 | 15.7 | 37.3 | 40.2 | 22.2 | 30.2 | 5.8 | 9.8 | 10.2 |
| 5 | 12.1 | 9.6 | 14.6 | 20.8 | 23.1 | 31.3 | 18.5 | 5.6 | 26.9 | 27.8 | 16.7 | 73 | 71.2 | 36.4 | 17.3 | 23.7 | 14.6 | 25.8 | 27.1 | 13.8 | 7.2 | 7.7 | 6 |
| 6 | 13.2 | 12.5 | 12.4 | 12.1 | 26.2 | 19.6 | 11.5 | 5.8 | 10.6 | 20.2 | 5.2 | 23.4 | 30.5 | 36.1 | 16.3 | 15.9 | 9.9 | 20.8 | 6 | 7.1 | 3.8 | 5.7 | 3.4 |
| 7 | 11.4 | 8 | 10.1 | 14 | 20.6 | 19.5 | 18.3 | 5.5 | 6.3 | 10.6 | 2.9 | 13.7 | 23.9 | 27.3 | 21.5 | 27.6 | 5.8 | 3.1 | 7.2 | 2.7 | 3.3 | 2.5 | 2.8 |
| 8 | 12.6 | 6.6 | 8.6 | 10.8 | 21.8 | 9 | 14.7 | 10.5 | 9.6 | 3.8 | 2.4 | 5.9 | 17.3 | 21.9 | 47.1 | 5.6 | 6 | 5 | 4.3 | 3.4 | 1.4 | 5.1 | 2.4 |
| 9 | 7.3 | 1.5 | 2.5 | 8.3 | 12.9 | 11.5 | 10.2 | 6.3 | 10.9 | 5.4 | 3.8 | 1.6 | 7.9 | 10.2 | 11.2 | 6.3 | 3.4 | 4.6 | 4 | 0.9 | 1.6 | 1.2 | 0.9 |
| 10 | 5.9 | 5.3 | 0.8 | 4 | 8.2 | 9 | 10 | 6.8 | 1.5 | 11 | 5.8 | 1.4 | 1.7 | 7.5 | 9.3 | 8.3 | 10.1 | 1.5 | 5.4 | 1 | 0.9 | 0.1 | 0.3 |
| 11 | 0 | 0.3 | 0.3 | 2.7 | 2.1 | 7 | 9.6 | 5.1 | 3.4 | 6.2 | 2.3 | 0.2 | 0.6 | 1.9 | 7.2 | 2.9 | 6.9 | 0.5 | 3.7 | 1.3 | 0.2 | 0.1 | 0.5 |
| 12 | 8.8 | 1.3 | 0.3 | 0.7 | 0.4 | 3.1 | 5.4 | 3 | 3.3 | 4.5 | 4.1 | 1.7 | 0.2 | 2.1 | 3.7 | 0.3 | 3.6 | 0.1 | 1 | 0.4 | 0.9 | 0.4 | 0 |
| 13 | 0.2 | 8.9 |  | 1.8 | 1.4 | 1.6 | 3.7 | 2.2 | 2.3 | 6.2 | 2.5 | 0.6 | 0.7 | 0.4 | 0.3 | 0.3 | 0.8 |  | 0.6 | 0 | 0.2 | 1.4 | 0 |
| 14 | 4.4 | 8 | 1.4 | 0.3 | 3.8 |  | 2 | 1.3 | 3.4 | 2.3 | 9.9 | 1 | 0.7 | 2.4 | 0.9 | 0.2 | 0.3 | 0.2 | 0 | 0.2 | 0.2 | 0.5 | 0.3 |
| 15+ |  |  |  | 5.1 | 4 | 12.2 | 5.8 | 2.7 | 4.7 | 8.5 | 9.6 | 0.8 |  | 1 | 6.1 | 1.1 | 0.5 |  | 0.1 | 0.1 |  |  | 0.3 |


| KG | WEI |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| 1 | 0.076 | 0.107 | 0.063 | 0.063 | 0.063 | 0.075 | 0.067 | 0.066 | 0.075 | 0.076 | 0.07 | 0.074 | 0.615 | 0.063 | 0.074 | 0.077 | 0.061 | 0.069 | 0.077 | 0.078 | 0.062 | 0.07 | 0.06 |
| 2 | 0.126 | 0.123 | 0.102 | 0.102 | 0.102 | 0.1 | 0.09 | 0.096 | 0.105 | 0.105 | 0.087 | 0.098 | 0.081 | 0.096 | 0.087 | 0.101 | 0.092 | 0.09 | 0.099 | 0.11 | 0.099 | 0.093 | 0.086 |
| 3 | 0.125 | 0.143 | 0.126 | 0.126 | 0.126 | 0.137 | 0.094 | 0.129 | 0.122 | 0.122 | 0.104 | 0.116 | 0.104 | 0.109 | 0.113 | 0.118 | 0.096 | 0.118 | 0.112 | 0.113 | 0.13 | 0.115 | 0.113 |
| 4 | 0.133 | 0.156 | 0.142 | 0.142 | 0.142 | 0.152 | 0.117 | 0.155 | 0.136 | 0.146 | 0.133 | 0.124 | 0.115 | 0.125 | 0.134 | 0.137 | 0.115 | 0.142 | 0.138 | 0.135 | 0.15 | 0.126 | 0.131 |
| 5 | 0.146 | 0.177 | 0.16 | 0.16 | 0.16 | 0.165 | 0.159 | 0.171 | 0.164 | 0.174 | 0.159 | 0.141 | 0.13 | 0.145 | 0.152 | 0.155 | 0.145 | 0.152 | 0.166 | 0.144 | 0.169 | 0.158 | 0.173 |
| 6 | 0.164 | 0.187 | 0.175 | 0.175 | 0.175 | 0.192 | 0.183 | 0.195 | 0.18 | 0.198 | 0.197 | 0.178 | 0.163 | 0.161 | 0.182 | 0.183 | 0.166 | 0.172 | 0.18 | 0.177 | 0.196 | 0.155 | 0.189 |
| 7 | 0.161 | 0.203 | 0.199 | 0.199 | 0.199 | 0.194 | 0.198 | 0.216 | 0.193 | 0.224 | 0.238 | 0.212 | 0.192 | 0.193 | 0.195 | 0.206 | 0.193 | 0.183 | 0.2 | 0.184 | 0.26 | 0.162 | 0.177 |
| 8 | 0.178 | 0.195 | 0.231 | 0.231 | 0.231 | 0.216 | 0.201 | 0.227 | 0.212 | 0.229 | 0.248 | 0.247 | 0.197 | 0.221 | 0.258 | 0.199 | 0.193 | 0.188 | 0.216 | 0.201 | 0.29 | 0.235 | 0.188 |
| 9 | 0.165 | 0.218 | 0.25 | 0.25 | 0.25 | 0.244 | 0.237 | 0.228 | 0.24 | 0.256 | 0.259 | 0.236 | 0.257 | 0.286 | 0.253 | 0.241 | 0.305 | 0.212 | 0.223 | 0.222 | 0.265 | 0.246 | 0.222 |
| 10 | 0.173 | 0.241 | 0.259 | 0.259 | 0.259 | 0.283 | 0.246 | 0.253 | 0.27 | 0.29 | 0.287 | 0.286 | 0.255 | 0.295 | 0.322 | 0.227 | 0.334 | 0.204 | 0.226 | 0.22 | 0.312 | 0.359 | 0.233 |
| 11 | 0.317 | 0.307 | 0.3 | 0.3 | 0.3 | 0.286 | 0.26 | 0.303 | 0.24 | 0.3 | 0.335 | 0.237 | 0.517 | 0.273 | 0.422 | 0.284 | 0.345 | 0.275 | 0.242 | 0.264 | 0.262 | 0.369 | 0.257 |
| 12 | 0.233 | 0.211 | 0.329 | 0.329 | 0.329 | 0.354 | 0.286 | 0.293 | 0.298 | 0.297 | 0.349 | 0.261 | 0.279 | 0.309 | 0.447 | 0.234 | 0.408 | 0.195 | 0.263 | 0.287 | 0.318 | 0.379 |  |
| 13 | 0.241 | 0.258 | 0.367 | 0.367 | 0.367 | 0.316 | 0.287 | 0.317 | 0.356 | 0.301 | 0.338 | 0.267 | 0.339 | 0.375 | 0.383 | 0.288 | 0.474 |  | 0.262 | 0.252 | 0.351 | 0.242 |  |
| 14 | 0.348 | 0.277 | 0.299 | 0.299 | 0.299 |  | 0.295 | 0.32 | 0.316 | 0.338 | 0.373 | 0.302 | 0.414 | 0.277 | 0.362 | 0.315 | 0.415 | 0.187 | 0.559 | 0.408 | 0.235 | 0.39 | 0.214 |
| 15+ | 0.348 | 0.277 | 0.36 | 0.36 | 0.36 | 0.35 | 0.336 | 0.389 | 0.353 | 0.402 | 0.375 | 0.404 |  | 0.389 | 0.46 | 0.351 | 0.475 |  | 0.339 | 0.273 |  | 0.378 | 0.26 |

$\qquad$ LENGTH

| AGE | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 1 | 19.2 | 19.2 | 19.2 | 19.2 | 19.2 | 19.1 | 19.5 | 19.4 | 20.3 | 19.8 | 18.1 | 20.1 | 19.9 | 20 | 20.3 | 20.8 | 19.2 | 19.9 | 20.9 | 20.4 | 19.8 | 20 | 19.1 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 22 | 22 | 22 | 22 | 22 | 21.5 | 21.5 | 21.7 | 22.3 | 22.2 | 21.5 | 22 | 20.8 | 21.6 | 21.6 | 22.6 | 21.7 | 21.7 | 22.4 | 22.9 | 22.9 | 22 | 21.3 |
| 3 | 23.5 | 23.5 | 23.5 | 23.5 | 23.5 | 23.9 | 21.9 | 23.8 | 23.7 | 23.6 | 22.9 | 23.4 | 22.5 | 23.2 | 23.2 | 23.9 | 23 | 23.5 | 23.5 | 23.6 | 24.6 | 23.6 | 23.3 |
| 4 | 24.8 | 24.8 | 24.8 | 24.8 | 24.8 | 24.9 | 23.4 | 25.4 | 24.6 | 25.2 | 24.7 | 24.1 | 23.6 | 24.1 | 24.6 | 25 | 24.5 | 25 | 25.3 | 24.8 | 25.8 | 24.8 | 24.1 |
| 5 | 25.5 | 25.5 | 25.5 | 25.5 | 25.5 | 26 | 26.7 | 26.3 | 26.2 | 26.6 | 25.9 | 25.4 | 24.4 | 25.6 | 25.8 | 25.7 | 25.9 | 25.7 | 27 | 25.4 | 26.6 | 26.4 | 26.7 |
| 6 | 26.4 | 26.4 | 26.4 | 26.4 | 26.4 | 27.6 | 27.5 | 27.4 | 27.3 | 27.5 | 27.7 | 27 | 26.6 | 26.3 | 27.2 | 27.1 | 27.6 | 27 | 27.1 | 27.3 | 28.2 | 26.1 | 27.5 |
| 7 | 27.2 | 27.2 | 27.2 | 27.2 | 27.2 | 28.1 | 28.1 | 28.6 | 28.2 | 28.8 | 29.8 | 28.6 | 27.8 | 28.1 | 28.1 | 28.3 | 27.7 | 27.1 | 28.3 | 27.5 | 30.4 | 27.5 | 27.5 |
| 8 | 29.2 | 29.2 | 29.2 | 29.2 | 29.2 | 28.6 | 28.5 | 29.3 | 29 | 29.2 | 30.4 | 29.8 | 28.1 | 28.8 | 30.6 | 28.4 | 27.8 | 27 | 28.9 | 28 | 31.7 | 30.2 | 28 |
| 9 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 29.9 | 29.8 | 29.4 | 29.9 | 30.4 | 30.8 | 30.8 | 30.1 | 31.2 | 31.1 | 30.2 | 31.9 | 28.6 | 29.2 | 28.8 | 30.5 | 30.5 | 29.1 |
| 10 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 31.2 | 30.2 | 30.3 | 30.9 | 31.4 | 31.8 | 31.5 | 31 | 31.8 | 32.5 | 30 | 32.5 | 28 | 29.5 | 29.2 | 32.5 | 34.7 | 29.5 |
| 11 | 30.6 | 30.6 | 30.6 | 30.6 | 30.6 | 31.5 | 30.7 | 31.4 | 30.7 | 31.9 | 33.8 | 31.2 | 39.5 | 31.6 | 35 | 32.2 | 33.2 | 30.1 | 30 | 30.7 | 31.5 | 35.2 | 31.1 |
| 12 | 32.1 | 32.1 | 32.1 | 32.1 | 32.1 | 33.6 | 32 | 31.6 | 31.9 | 31.7 | 35.6 | 30.8 | 31.5 | 32.2 | 35.3 | 30.8 | 34.6 | 27.5 | 30.4 | 30.6 | 32.3 | 35.5 |  |
| 13 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 33.3 | 31.7 | 32.4 | 32.8 | 31.9 | 34 | 32.1 | 33.4 | 33.9 | 34 | 31.8 | 36.4 |  | 32.1 | 30 | 32.5 | 31.5 |  |
| 14 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 |  | 32.1 | 32.4 | 32.5 | 33 | 34.4 | 32.5 | 34.5 | 32.3 | 34.2 | 33 | 36 | 27.5 | 38.5 | 36 | 30.5 | 36.1 | 30.5 |
| $15+$ | 32.5 | 32.5 | 32.5 | 32.5 | 32.5 | 33.8 | 33.4 | 34.3 | 33.6 | 34.8 | 35.2 | 35.3 |  | 35.1 | 36.1 | 34.5 | 36.9 |  | 34.2 | 32.5 | 36.1 | 31.5 |  |

Table 6.3.3. North Sea Horse Mackerel stock. Mean weight at age (kg) in the catch by quarter and area in 2017.

| Number/1000 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.06 | 0.06 | 0.06 | 0.06 | 0 | 0.06 |
| 2 | 0.086 | 0.086 | 0.086 | 0.086 | 0.058 | 0.081 |
| 3 | 0.114 | 0.114 | 0.114 | 0.114 | 0.087 | 0.108 |
| 4 | 0.13 | 0.13 | 0.13 | 0.13 | 0.112 | 0.127 |
| 5 | 0.172 | 0.172 | 0.172 | 0.172 | 0.151 | 0.168 |
| 6 | 0.19 | 0.19 | 0.19 | 0.19 | 0.159 | 0.184 |
| 7 | 0.175 | 0.175 | 0.175 | 0.175 | 0.163 | 0.173 |
| 8 | 0.188 | 0.188 | 0.188 | 0.188 | 0.178 | 0.186 |
| 9 | 0.222 | 0.222 | 0.222 | 0.222 | 0.225 | 0.223 |
| 10 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 11 | 0.257 | 0.257 | 0.257 | 0.257 | 0.277 | 0.261 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 |
| 15 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 |
| 2Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 |
| 2 | 0.086 | 0.086 | 0.086 | 0.086 | 0.086 | 0.086 |
| 3 | 0.114 | 0.114 | 0.114 | 0.114 | 0.114 | 0.114 |
| 4 | 0.13 | 0.13 | 0.13 | 0.13 | 0.13 | 0.13 |
| 5 | 0.172 | 0.172 | 0.172 | 0.172 | 0.172 | 0.172 |
| 6 | 0.19 | 0.19 | 0.19 | 0.19 | 0.19 | 0.19 |
| 7 | 0.175 | 0.175 | 0.175 | 0.175 | 0.175 | 0.175 |
| 8 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 |
| 9 | 0.222 | 0.222 | 0.222 | 0.222 | 0.222 | 0.222 |
| 10 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 |
| 11 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 |


| 15 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0.06 | 0.06 | 0.06 | 0.06 |
| 2 | 0 | 0 | 0.086 | 0.086 | 0.086 | 0.086 |
| 3 | 0 | 0 | 0.114 | 0.114 | 0.114 | 0.114 |
| 4 | 0 | 0 | 0.13 | 0.13 | 0.13 | 0.13 |
| 5 | 0 | 0 | 0.172 | 0.172 | 0.172 | 0.172 |
| 6 | 0 | 0 | 0.19 | 0.19 | 0.19 | 0.19 |
| 7 | 0 | 0 | 0.175 | 0.175 | 0.175 | 0.175 |
| 8 | 0 | 0 | 0.188 | 0.188 | 0.188 | 0.188 |
| 9 | 0 | 0 | 0.222 | 0.222 | 0.222 | 0.222 |
| 10 | 0 | 0 | 0.233 | 0.233 | 0.233 | 0.233 |
| 11 | 0 | 0 | 0.257 | 0.257 | 0.257 | 0.257 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0.214 | 0.214 | 0.214 | 0.214 |
| 15 | 0 | 0 | 0.26 | 0.26 | 0.26 | 0.26 |
| 4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0.06 | 0.06 | 0.059 | 0.074 |
| 2 | 0 | 0 | 0.086 | 0.092 | 0.088 | 0.089 |
| 3 | 0 | 0 | 0.114 | 0.115 | 0.117 | 0.115 |
| 4 | 0 | 0 | 0.13 | 0.141 | 0.148 | 0.14 |
| 5 | 0 | 0 | 0.172 | 0.18 | 0.198 | 0.183 |
| 6 | 0 | 0 | 0.19 | 0 | 0.206 | 0.132 |
| 7 | 0 | 0 | 0.175 | 0 | 0.215 | 0.13 |
| 8 | 0 | 0 | 0.188 | 0 | 0.208 | 0.132 |
| 9 | 0 | 0 | 0.222 | 0 | 0.219 | 0.147 |
| 10 | 0 | 0 | 0.233 | 0 | 0 | 0.078 |
| 11 | 0 | 0 | 0.257 | 0 | 0.234 | 0.164 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0.214 | 0 | 0 | 0.071 |
| 15 | 0 | 0 | 0.26 | 0 | 0 | 0.087 |


| 1-4Q |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.06 | 0.06 | 0.06 | 0.06 | 0.059 | 0.075 |
| 2 | 0.086 | 0.086 | 0.086 | 0.088 | 0.081 | 0.087 |
| 3 | 0.114 | 0.114 | 0.114 | 0.114 | 0.109 | 0.117 |
| 4 | 0.13 | 0.13 | 0.13 | 0.133 | 0.131 | 0.135 |
| 5 | 0.172 | 0.172 | 0.172 | 0.174 | 0.174 | 0.167 |
| 6 | 0.19 | 0.19 | 0.19 | 0.19 | 0.187 | 0.178 |
| 7 | 0.175 | 0.175 | 0.175 | 0.175 | 0.182 | 0.179 |
| 8 | 0.188 | 0.188 | 0.188 | 0.188 | 0.191 | 0.257 |
| 9 | 0.222 | 0.222 | 0.222 | 0.222 | 0.222 | 0.291 |
| 10 | 0.233 | 0.233 | 0.233 | 0.233 | 0.233 | 0.359 |
| 11 | 0.257 | 0.257 | 0.257 | 0.257 | 0.255 | 0.369 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.379 |
| 15 | 0.26 | 0.26 | 0.26 | 0.26 | 0.26 | 0.37 |

Table 6.3.4. North Sea Horse Mackerel stock. Mean length (cm) at age in the catch by quarter and area in 2017.

| Number/1000 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 19.1 | 19.1 | 19.1 | 19.1 | 0 | 19.1 |
| 2 | 21.3 | 21.3 | 21.3 | 21.3 | 18 | 20.7 |
| 3 | 23.4 | 23.4 | 23.4 | 23.4 | 21.5 | 23 |
| 4 | 24.1 | 24.1 | 24.1 | 24.1 | 22.9 | 23.8 |
| 5 | 26.6 | 26.6 | 26.6 | 26.6 | 25.7 | 26.4 |
| 6 | 27.5 | 27.5 | 27.5 | 27.5 | 26.3 | 27.3 |
| 7 | 27.4 | 27.4 | 27.4 | 27.4 | 26.8 | 27.3 |
| 8 | 28 | 28 | 28 | 28 | 27.8 | 28 |
| 9 | 29.1 | 29.1 | 29.1 | 29.1 | 29.6 | 29.2 |
| 10 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 |
| 11 | 31.1 | 31.1 | 31.1 | 31.1 | 32.5 | 31.4 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 |
| 15 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 |
| 2Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 19.1 | 19.1 | 19.1 | 19.1 | 19.1 | 19.1 |
| 2 | 21.3 | 21.3 | 21.3 | 21.3 | 21.3 | 21.3 |
| 3 | 23.4 | 23.4 | 23.4 | 23.4 | 23.4 | 23.4 |
| 4 | 24.1 | 24.1 | 24.1 | 24.1 | 24.1 | 24.1 |
| 5 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 |
| 6 | 27.5 | 27.5 | 27.5 | 27.5 | 27.5 | 27.5 |
| 7 | 27.4 | 27.4 | 27.4 | 27.4 | 27.4 | 27.4 |
| 8 | 28 | 28 | 28 | 28 | 28 | 28 |
| 9 | 29.1 | 29.1 | 29.1 | 29.1 | 29.1 | 29.1 |
| 10 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 |
| 11 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 |


| 15 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 19.1 | 19.1 | 19.1 | 19.1 |
| 2 | 0 | 0 | 21.3 | 21.3 | 21.3 | 21.3 |
| 3 | 0 | 0 | 23.4 | 23.4 | 23.4 | 23.4 |
| 4 | 0 | 0 | 24.1 | 24.1 | 24.1 | 24.1 |
| 5 | 0 | 0 | 26.6 | 26.6 | 26.6 | 26.6 |
| 6 | 0 | 0 | 27.5 | 27.5 | 27.5 | 27.5 |
| 7 | 0 | 0 | 27.4 | 27.4 | 27.4 | 27.4 |
| 8 | 0 | 0 | 28 | 28 | 28 | 28 |
| 9 | 0 | 0 | 29.1 | 29.1 | 29.1 | 29.1 |
| 10 | 0 | 0 | 29.5 | 29.5 | 29.5 | 29.5 |
| 11 | 0 | 0 | 31.1 | 31.1 | 31.1 | 31.1 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 30.5 | 30.5 | 30.5 | 30.5 |
| 15 | 0 | 0 | 31.5 | 31.5 | 31.5 | 31.5 |
| 4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 19.1 | 19.2 | 19.1 | 19.1 |
| 2 | 0 | 0 | 21.3 | 21.9 | 21.6 | 21.6 |
| 3 | 0 | 0 | 23.4 | 23.4 | 23.6 | 23.5 |
| 4 | 0 | 0 | 24.1 | 24.8 | 25.3 | 24.7 |
| 5 | 0 | 0 | 26.6 | 28.5 | 27.5 | 27.5 |
| 6 | 0 | 0 | 27.5 | 0 | 28.1 | 18.5 |
| 7 | 0 | 0 | 27.4 | 0 | 28.7 | 18.7 |
| 8 | 0 | 0 | 28 | 0 | 28.5 | 18.8 |
| 9 | 0 | 0 | 29.1 | 0 | 28.4 | 19.2 |
| 10 | 0 | 0 | 29.5 | 0 | 0 | 9.8 |
| 11 | 0 | 0 | 31.1 | 0 | 29.5 | 20.2 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 30.5 | 0 | 0 | 10.2 |
| 15 | 0 | 0 | 31.5 | 0 | 0 | 10.5 |


| 1-4Q |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 19.1 | 19.1 | 19.1 | 19.1 | 19.1 | 19.1 |
| 2 | 21.3 | 21.3 | 21.3 | 21.5 | 20.7 | 21.3 |
| 3 | 23.4 | 23.4 | 23.4 | 23.4 | 23 | 23.3 |
| 4 | 24.1 | 24.1 | 24.1 | 24.3 | 24.1 | 24.1 |
| 5 | 26.6 | 26.6 | 26.6 | 27.1 | 26.6 | 26.7 |
| 6 | 27.5 | 27.5 | 27.5 | 27.5 | 27.4 | 27.5 |
| 7 | 27.4 | 27.4 | 27.4 | 27.4 | 27.6 | 27.5 |
| 8 | 28 | 28 | 28 | 28 | 28.1 | 28 |
| 9 | 29.1 | 29.1 | 29.1 | 29.1 | 29 | 29.1 |
| 10 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 | 29.5 |
| 11 | 31.1 | 31.1 | 31.1 | 31.1 | 31 | 31.1 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 |
| 15 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 | 31.5 |

Table 6.4.1. North Sea Horse Mackerel. cpue Indices of abundance (individuals/hour) for juvenile $(<20 \mathrm{~cm})$ and exploitable $(>20 \mathrm{~cm})$ substocks, estimated as a combined index for the NS-IBTS Q3 (North Sea only, no 27.7.d included) and the Channel Ground Fish Survey in Q4 (CGFS, 27.7.d). The survey indices are derived from the prediction of a hurdle model fit to data over the period 1992-2017.

|  | Juvenile substock (<20cm) |  |  | Exploitable substock ( $>20 \mathrm{~cm}$ ) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Index | CI_low | CI_high | Index | CI_low | CI_high |
| 1992 | 4865 | 2293 | 9237 | 1498 | 663 | 2915 |
| 1993 | 1917 | 959 | 3415 | 565 | 291 | 974 |
| 1994 | 3288 | 1593 | 6067 | 1322 | 635 | 2368 |
| 1995 | 2232 | 1107 | 4115 | 1621 | 669 | 3401 |
| 1996 | 1178 | 447 | 2480 | 1080 | 482 | 1987 |
| 1997 | 3350 | 1516 | 6253 | 714 | 336 | 1286 |
| 1998 | 858 | 414 | 1573 | 436 | 201 | 806 |
| 1999 | 1475 | 794 | 2433 | 517 | 257 | 905 |
| 2000 | 1139 | 516 | 2333 | 289 | 137 | 570 |
| 2001 | 3431 | 1580 | 7437 | 508 | 245 | 916 |
| 2002 | 2999 | 1515 | 5386 | 501 | 240 | 937 |
| 2003 | 2116 | 1190 | 3499 | 381 | 179 | 726 |
| 2004 | 1064 | 559 | 1844 | 428 | 199 | 754 |
| 2005 | 987 | 530 | 1727 | 759 | 366 | 1370 |
| 2006 | 502 | 271 | 880 | 834 | 422 | 1523 |
| 2007 | 665 | 375 | 1107 | 411 | 197 | 787 |
| 2008 | 394 | 224 | 695 | 209 | 101 | 458 |
| 2009 | 758 | 416 | 1265 | 104 | 47 | 212 |
| 2010 | 1611 | 863 | 2981 | 234 | 106 | 459 |
| 2011 | 569 | 317 | 1091 | 282 | 136 | 583 |
| 2012 | 354 | 189 | 705 | 185 | 93 | 437 |
| 2013 | 1062 | 572 | 1882 | 146 | 64 | 335 |
| 2014 | 1609 | 909 | 2747 | 430 | 193 | 876 |
| 2015 | 2257 | 1220 | 4527 | 580 | 261 | 1146 |
| 2016 | 1752 | 959 | 2976 | 803 | 376 | 1557 |
| 2017 | 973 | 505 | 1714 | 131 | 54 | 282 |

### 6.9 Figures

TAC and catch uptake. Year 2017


Figure 6.2.1. North Sea horse mackerel. Utilisation of quota by country.


Figure 6.2.2 North Sea Horse Mackerel. North Sea horse mackerel. Catch by ICES Division for 20002017.


Figure 6.2.3.- North Sea Horse Mackerel. Total catch (in tonnes) by area, quarter, catch category and country. BMS landing refers to landings below minimum legal size.

North Sea Stock: Catch by division


Figure 6.3.1.- North Sea Horse Mackerel. Proportion of NSHM total catch per year and station that have been sampled.

NSHM: catch at age ( N ; observed) all areas


Figure 6.3.2.- North Sea horse mackerel age distribution in the catch for 1995-2017. The area of bubbles is proportional to the catch number. Note that age 15 is a plus group.


NSHM: catch at age ( N ; observed) out of 27.7.d


Figure 6.3.3. North Sea horse mackerel. Bubbleplots of age distribution in the catch by area for 19982017 for area 7.d (upper panel) and out of 7.d (bottom panel). The area of bubbles is proportional to the total catch number for the stock. Note that age 15 is a plus group.


Figure 6.3.4. North Sea horse mackerel. Mean weight at age in commercial catches over the period 2000-2017

## Mean length at age (cm)



Figure 6.3.5. North Sea horse mackerel. Mean length at age in commercial catches over the period 2000-2017.


Figure 6.4.1. North Sea Horse Mackerel. Catch curves for the 1994 to 2007 cohorts, ages from 3 to 14. Values plotted are the $\log ($ catch ) values for each cohort in each year. The negative slope of these curves estimates total mortality $(Z)$ in the cohort.

Total mortality by cohort


Figure 6.4.2. North Sea Horse Mackerel. Total mortality by cohort ( $Z$ ) estimated from the negative gradients of the 1992-2006 cohort catch curves (Figure 6.4.1).


Figure 6.4.3.- North Sea horse mackerel. ICES rectangles selected in 2013 and currently used by the working group.


Figure 6.4.4.- North Sea horse mackerel. Size distribution of North Sea horse mackerel catches during the inter-calibration exercise conducted in 2014 between the RV Gwen Drez (red bars) and Thalassa (blue bars).


Figure 6.4.5. North Sea horse mackerel. cpue by depth for the CGFS survey from 1992 to 2017.


Figure 6.4.6. North Sea Horse Mackerel. Combined cpue survey index (indiv/hour) derived from the hurdle model fit to the IBTS survey in the North Sea (4.bc) and the CGFS survey in the English channel. Top: Juvenile substock ( $<20 \mathrm{~cm}$ ); Bottom: exploitable substock ( $>20 \mathrm{~cm}$ ). The abundance index values are presented as number of individuals per hours. The confidence interval is determined by bootstrap resampling of Pearson residuals with 1000 iterations.


Figure 6.4.7.- North Sea horse mackerel. Proportion of hauls with zero catch for the exploitable ( $\mathbf{~} 20 \mathrm{~cm}$ ) and juvenile ( $<20 \mathrm{~cm}$ ) substocks in the NS-IBTS (pink dotted lines) and the CGFS (blue dotted lines).

>20cm substock


Figure 6.4.8. North Sea Horse Mackerel. Mean CPUE survey index (indiv/hour) obtained from the hurdle model fit to the IBTS survey in the North Sea (4.bc) and the CGFS survey in the English channel. Top: Juvenile substock ( $<20 \mathrm{~cm}$ ); Bottom: exploitable substock ( $>20 \mathrm{~cm}$ ). The abundance index values are presented as number of individuals per hours.


Figure 6.4.9. North Sea horse mackerel. Relative occurrence by length for the period 2012-2017 in the CGFS.


Figure 6.4.10. North Sea horse mackerel. Relative occurrence by length for the period 2012-2017 in the IBTS.


Figure 6.4.11.- Length distribution, as well as the estimated parameters Lc, Lmean, $\mathrm{Lf}=\mathrm{m}$ for the Dutch fleet in 2016 and 2017.

## 7 Western Horse Mackerel -in Subarea 8 and divisions 2.a, 3.a (Western Part), 4.a, 5.b, 6.a, 7.a-c and7.e-k

### 7.1 ICES advice applicable to 2017and 2018

Since 2011, the TACs cover areas in line with the distribution areas of the stock.
For 2017 the TAC set in EU waters (EU 2017/127) was the following:

| Areas in EU Waters | TAC 2017 | Stocks fished in this area |
| :---: | :---: | :---: |
| - 2.a, 4.a, 5.b, Subareas 6, 7.a-c, 7.e-k, 8.abde, 12, 14 | 82229 t | Western stock \& North Sea stock in 4.a 1-2 quarters |
| 4.b,c, 7.d | 14 697t | North Sea stocks |
| Division 8.c | 13271 t | Western stock |

For 2018 the TAC set in EU waters (EU 2018/120) was the following:

| Areas in EU Waters | TAC 2018 | Stocks fished in this area |
| :---: | :---: | :---: |
| - 2.a, 4.a, 5.b, Subareas 6, 7.a-c, 7.e-k, 8.abde, 12, 14 | 99470 t | Western stock \& North Sea stock in 4.a 1-2 quarters |
| 4.b,c, 7.d | 12629 t | North Sea stocks |
| Division 8.c | 16000 t | Western stock |

The TAC for the western stock should apply to the distribution area of western horse mackerel as follows:

All Quarters: 2.a, 5.b, 6.a, 7.a-c, 7.e-k, 8.a-e
Quarters 3\&4: 3.a (west), 4.a
The TAC for the North Sea stock should apply to the distribution area of North Sea horse mackerel as follows:

All Quarters: $\quad 3 . a$ (east), 4.b-c, 7.d
Quarters 1\&2: 3.a (west), 4.a
In 2017 ICES advised on the basis of MSY approach that Western horse mackerel catches in 2018 should be no more than 117070 tonnes. The Western horse mackerel TAC for 2018 is 117070 tonnes, the TAC for EU waters only is 115470 tonnes. The TAC should apply to the total distribution area of this stock. The EU horse mackerel catches in Division 3.a are taken outside the horse mackerel TACs.

### 7.1.1 The fishery in 2017

Information on the development of the fisheries by quarter and division is shown in Tables 5.1.1 and 5.1.2 and in Figures 5.1.1.a-d. The total catch allocated to western horse mackerel in 2017 was 82961 t which is 15850 t less than in 2016 and in line with ICES advice. The catches of horse mackerel by country and area are shown in Tables 7.1.1.15 while the catches by quarter since 2000 are shown in Figure 7.1.1.1

### 7.1.2 Estimates of discards

Discard data are available since 2000 for few countries. Until 2013 however the estimates available are considered an underestimation of the overall amount. (Figure 7.1.2.1)

In 2017 most countries have presented discard information. Countries that reported discard for horse mackerel were Spain, Denmark, France, Sweden, UK (England + Wales) and Scotland. 2017 discard for Germany, Ireland, Netherland, Norway and Faroe Island is considered equal to zero. Discard rate for western horse mackerel was equal to 3926 tonnes, equal to $4.3 \%$ in weight of the total catches.
Discard data are included in the assessment as part of the total catches.
Length frequency distributions of discard were provided by Spain and France.

### 7.1.3 Stock description and management units

The western horse mackerel stock spawns in the Bay of Biscay, and in UK and Irish waters. After spawning, parts of the stock migrate northwards into the Norwegian Sea and the North Sea, where they are fished in the third and fourth quarter. The stock is distributed in Divisions 2.a, 5.b, 3.a, 4.a, 6.a, 7.a-c, 7.e-k and 8.a-e. The stock is caught in these areas following the yearly distribution described in Section 5.3 (Figures 7.1.3.12). The western stock is considered a management unit and advised accordingly. At present there are no international agreed management measures. EU regulates the fishery by TAC. This TAC is now set in accordance with the distribution of the stock although catches in 3.a are taken outside the TAC.

### 7.2 Scientific data

### 7.2.1 Egg survey estimates

Last egg survey was carried out in the western and southern spawning areas in 2016 and a presentation with the final results were given during the WGWIDE meeting by the survey coordinator in 2017 (ICES, 2017a).

The time-series of egg production estimates for western horse mackerel is presented in Table 7.2.1.1 and shown in figure 7.2.1.1

The ICES Working Group on Mackerel and Horse Mackerel Egg Surveys (WGMEGS) (ICES 2018) met in Dublin on April 2018, to plan the 2019 Mackerel and Horse Mackerel Egg Survey for the Western horse mackerel stock.

### 7.2.2 Other surveys for western horse mackerel

Bottom-trawl surveys
An update bottom-trawl survey index for recruitment was available for the 2017 assessment: the index is based on IBTS surveys conducted by Ireland, France and Scotland covering the main distribution of the stock (Bay of Biscay, Celtic Sea, West of Ireland and West of Scotland) from 2003 to 2017, and uses a Bayesian Delta-GLMM for the calculation of an index of juvenile abundance based on catch rates (ICES 2017b). The update index is shown in figure 7.2.2.1 (plot in the middle). The 2017 data point is highly uncertain due to the very little coverage of the French survey: the French research vessel had technical issue and could therefore only cover less than $1 / 3$ of the stations usually sampled. Despite this huge uncertainty, the 2017 data point suggests a very strong recruitment to be expected next year. This perception is confirmed by the
presence of numerous small fish in the 2017 catch data. The overall trend suggests an increase in recruitment from 2013, being the 2017 estimates one of the highest of the time-series; anecdotal information from Spanish fisheries independent surveys confirms the good incoming recruitment.
Further information on how the recruitment index is estimated can be found in the stock annex, in ICES (2008/ACOM:13), ICES (2009/RMC:04) and in ICES (2017b).
Acoustic surveys
In the Bay of Biscay two coordinated acoustic surveys are taking place at the spring time, PELGAS (Ifremer-France) and PELACUS (IEO-Spain)
PELACUS 0318 was carried out on board RV Miguel Oliver from 25 $5^{\text {th }}$ March till $18^{\text {th }}$ April. The methodology was similar to that of the previous surveys (. This year the surveyed area was steamed clockwise from the inner part of the Bay of Biscay to the Spanish-Portuguese border, thus contrary to the normal procedure (see Carrera et al. WD03; Carrera et al. WD04) for further details). Weather conditions were adverse, with a continuous low pressure fronts with dominant SW/W winds and swell of about 4 m height, resulted in an important haline front all around the surveyed area due to the river run-offs, and a poleward current with clear influence up to $6^{\circ} 30^{\prime}$ W (Galician waters). Fish spatial distribution was rather coastal for all species but blue whiting and pearlside.

Horse mackerel mainly occurred at the inner part of the Bay of Biscay and also in 9a (southern component), but in general the density was low (figure 7.2.2.2). This resulted in an estimation of 9 thousand tonnes, a clear decreasing trend since 2015 when biomass peaked at 67 thousand tonnes. Nevertheless it should be noted that age 1 and 2 accounted up to $88 \%$ of the abundance estimates. Both year classes were shown higher figures than those estimated in 2017 (figure 7.2.2.3).

In French waters, PELGAS took place in May (e.g. one month later), steamed northwards (Duhamel et al., WD11). Main concentrations were located between Garonne and Arcachon areas (figure 7.2.2.4), and well distributed throughout the surveyed area. This resulted in 92 thousand tonnes, higher than that estimated last year (figure 7.2.2.5), but still below the average time-series estimate (2000-18). On the other hand, as in the case of PELACUS, the bulk of the estimation was composed by young fish.

### 7.2.3 Effort and catch per unit effort

No new information was presented on effort and catch per unit effort. Further information can be found in the stock annex.

### 7.2.4 Catch in numbers

In 2017, the Netherlands (4.a, 6.a, 7.b,c,e,h,j ), Norway (4.a, 3.a), Ireland (6.a and 7.b,g), Germany (6.a, 7.b.e) and Spain (8.abc) provided catch in numbers at age. The catch sampled for age readings in 2017 covered $63 \%$, in 2016 covered $82 \%$ and in 2015 covered 69\%. (Figure 7.2.4.1). In addition France (4.a, 6a, 7.e,g,h, 8.a,b), England (7.e,f,g) and Scotland (6.a) provided catch in number at length.
The total annual and quarterly catches in number for western horse mackerel in 2017 are shown in Table 7.2.4.1. The sampling intensity is discussed in Section 5.9.
The catch-at-age matrix is given in Table 7.2.4.2, and illustrated in Figure 7.2.4.2. It shows the dominance of the 1982 year class in the catches since 1984 until it entered the plus group in 1997. Since 2002, the 2001 year class, which has now entered the plus
group in 2016, has been caught in considerable numbers. The 2008 year class can be followed in the catch data suggesting it was stronger than other year classes subsequent to the 2001.

In addition, Germany and Spain provided the Age Length Keys (ALK) which were used in 2017 to convert catch at length into catch-at-age.

### 7.2.5 Mean length at age and mean weight at age

## Mean length at age and mean weight at age in the catches

The mean weight and mean length at age in the catches by area, and by quarter in 2017 are shown in Tables 7.2.5.1 and 7.2.5.2. Weight at age time-series is shown in Figure 7.2.5.1.

Mean weight at age in the stock
Mean weight-at-age in the stock is presented in Table 7.2.5.3. Further information can be found in the stock annex.

### 7.2.6 Maturity ogive

Maturity-at-age is presented in Table 7.2.6.1. In the assessment model a constant logistic function was used (figure 7.2.6.1). Further information can be found in the stock annex.

### 7.2.7 Natural mortality

A fixed natural mortality of 0.15.year ${ }^{-1}$ is assumed for all ages and years in the assessment. Further information can be found in the stock annex.

### 7.2.8 Fecundity data

Potential fecundity data ( $10^{6} \mathrm{eggs}$ ) per kg spawning females are available for the years 1987, 1992, 1995, 1998, 2000, 2001: the data are presented in table 7.2.8.1 but were not used in the assessment model. In the assessment the fecundity is modelled as linear eggs $/ \mathrm{kg}$ on body weight. Further information can be found in the stock annex.

### 7.2.9 Information from the fishing industry

The fishing industry in conjunction with the Pelagic AC (PELAC) has been working actively on a number issues namely a large-scale genetics project on stock identification, development of a management strategy with the scientists and number of voluntary industry measure to protect juveniles.

The genetic work is now close to being finalized. Samples have been collected during the years 2015-2017 from area between Mauritania and the Northern North Sea. DNA has been extracted from the samples and are currently being worked up. It is expected that the genetic analysis will be finished in the first half of 2019.

The Irish and Dutch fishing industry reported good catches of horse mackerel catches southwest of Ireland (division 7.j) and west of Scotland (Division 6.a) during the first months of 2018, also including bigger sizes of horse mackerel (e.g. WGWIDE WD17).

### 7.2.10 Data exploration

The length frequency distributions of the catches for the whole fleet included in the model are shown in Figure 7.2.10.1. The length distributions available for 2015-2017
show a considerable amount of very small fish, mostly driven by the Spanish catches. Length frequency distribution from discards was analysed alongside the length frequency distribution from the landings. The huge numbers of small individuals from the discards have a strong impact on the overall LFD of the catches. These data were not available at the benchmark and to include those in the assessment model would require major changes in the modelling structure: for this reason were only used in the explorative analysis.

Within-cohort consistency of the catch-at-age matrix is investigated in Figure 7.2.10.2: this shows that the catch-at-age data contains information on year-class strength that could form the basis for an age-structured model.

The numbers at age in the catch by decade show a trend towards younger individuals when moving from the beginning of the time-series towards the end (Figure 7.2.10.3).

The indices of abundance used in the assessment cover different areas and represents different part of the stock. Negative correlations between indices that should represent the same portion of the population might cause problem in the fitting. The correlation between time-series was therefore estimated and presented in figure 7.2.10.4. Given the fact that the IBTS index is a recruitment index and that most of the juveniles are located in the area partially covered by the PELACUS survey, a higher correlation is expected. On the other hand, the egg survey should represent the adult portion of the stock: since no stock-recruitment relationship has been observed for the western horse mackerel stock, it is not surprising that the trajectory of this SSB index differs from the other two.

### 7.2.11 Assessment model, diagnostics

A one fleet, one sex, one area stock synthesis model (SS; Stock Synthesis v3.30; Methot, 2011) is used for the assessment of western horse mackerel stock in the Northeast Atlantic. A description of the model can be found in the stock annex. The assessment is presented as an update to the benchmark, and sees the inclusion of the 2017 estimates for the three surveys used, the 2017 length frequency distribution from the catches and the PELACUS survey and the 2017 total catch and conditional ALKs.

Fits to the available data are given in Figure 7.2.11.1, and model estimates with associated precision in Figure 7.2.11.2. Model estimates and residual patterns are similar to those presented in the benchmark (ICES, 2017b) and remain unchanged from last year assessment for almost all variables. Recruitment estimates for the years 2015-2017 were revised downwards with consequent degradation of the fitting to the IBTS recruitment index: this revision was caused by the model expecting a larger number of bigger/older fish in the catches that didn't occur. The model fitting to the most recent length frequency distributions and the conditional ALKs remain not optimal, due to changes in the overall pattern of the catches with a significant increase of smaller fish compared to the past.

Retrospective plots are shown for 10 years. Major rescaling of the estimates are observed in correspondence of the availability of a new egg survey data point. The inclusion of the 2016 length frequency distribution also caused a major deviation from the previous year assessment. The 2017 assessment shows only a minor revision upwards of the SSB, and a minor revision downwards of recruitment and F.

### 7.3 State of the Stock

### 7.3.1 Stock assessment

The SS model with new length and age data from the commercial fleet, and the 2017 information from the 3 surveys available, is presented as the final assessment model. Stock numbers-at-age and fishing mortality-at-age are given in Tables 7.3.1.1 and 7.3.1.2, and a stock-summary is provided in Table 7.3.1.3, and illustrated in Figure 7.3.1.1. and 7.3.1.2. SSB peaked in 1988 following the very strong 1982 year class. Subsequently SSB slowly declined till the second lowest value of the time-series in 2003 and then recovered again following the moderate-to-strong year class of 2001 (a third of the size of the 1982 year class). Year classes following 2001 have been weak: 2011, 2010 and 2013 recruitment in particular has been estimated as the lowest values in the time-series together with the 1983. The 2008 year class has been estimated to be fairly strong. Recruitment estimates for 2015-2017 are the highest observed since 2008 and are higher than the geometric mean estimated over the years 1983-2017. SSB in 2017 is estimated as the lowest in the time-series. Fishing mortality has been increasing since 2007 as a result of increasing catches and decreasing biomass as the 2001 year class was reduced. Since 2012 the F has been decreasing, dropping to low values in 2015-2017 due to lower catches and a reduced proportion of the adult population in the exploited stock.

### 7.4 Short-term forecast

A deterministic short-term forecast was conducted using the ' fwd()$^{\prime}$ ' method in FLR (Flash R add-on package).

## Input

Table 7.4.1. lists the input data for the short-term predictions. Weight at age in the stock and weight at age in the catch are equal to the year invariant weight at age function used in the stock synthesis model. Exploitation pattern is based on F in 2017 and is the average of ages 1 to 10 . Natural mortality is assumed to be 0.15 across all ages. The proportion mature for this stock has a logistic form with fully mature individuals at age 4 as used in the assessment model. As with last year, the expected landings for the intermediate year were set to the level that corresponds to the 2017 TAC in EU waters, equal to 115470 t . Note that -despite the plus group in the catch being equal to $15+$ - the true population in SS model is set to arrive up to age 20 (as from literature) and is therefore estimated accordingly.

Output
A range of predicted catch and SSB options from the short-term forecast are presented in Table 7.4.2.

### 7.5 Uncertainties in the assessment and forecast

Despite the increased amount of data used and information available to the stock assessment, the model still suffers of retrospective pattern whenever a new year of data is included. This year rescaling is however small compared to past assessment and it influenced by the 2017 length frequency distribution, which is skewed towards a higher amount of small specimen.

The fitting to the fishery independent indices remains good for two of the three surveys used: a degradation of the fitting to the IBTS recruitment index is observed, but the
estimates remain within the confidence intervals provided. The fit the acoustic index remains poor.

The change in selectivity, which is detected from both the length and the age composition of the catch data, is not entirely picked up from the model. In general, SS tend to overestimate the mean age of the last decade, as well as underestimate the presence of small individuals. The selectivity issue should be further investigated and somehow addressed: for example, it is not clear whether the high presence of small specimen in the landings data is due to the inclusion of BMS individuals in the overall catch instead of having it as discard (the discard ban applies since 2015) or if this is due to an effective change in selectivity (i.e. catchability of the gear and availability of the stock).

The model fix the realised fecundity with a constant number of eggs $/ \mathrm{kg}$ independently of the individual weight. However, western horse mackerel is known to be an indeterminate spawner, which implies this relationship being not appropriate when it comes to the use of an egg survey as index of spawning biomass. During the benchmark it was attempted to estimate the parameters relative to fecundity, but the information provided were not enough. The inclusion of this feature, whenever appropriate data will become available, would help to improve the reliability of the assessment.

The assumed value for $M$ should be investigated. However, there is no data available (such as tagging) that could assist in estimating M more accurately. Nevertheless, total mortality appears to be low, given the persistence of the 1982 year class in the catch data.

In general Stock Synthesis tends to underestimate the uncertainty of the main variables: in the present case, the estimated uncertainty, despite being low, remains higher than the yearly fluctuations; it is therefore considered reasonable.

The assessment, as was built at the benchmark, has now a fairly good amount of information for the estimates of recruitment, which is also informed by the strong, occasional year classes observed in the catch. On the contrary, the SSB is informed only by the triennial egg survey and by the acoustic survey (that however covers only a small part of the stock distribution and targets the smallest fraction of the population, has a really low weight in the model and is really noisy): a new index for the spawning biomass would therefore be beneficial for the future stability of this assessment. The development of a SSB index from the IBTS survey as well as merging the information available from the Pelacus and the Pelgas acoustic survey in the Bay of Biscay should be pursued.

### 7.6 Comparison with previous assessment and forecast

A comparison of the update assessment (with and without the inclusion of the 2016 data) with the benchmark assessment is shown in Figure 7.2.11.3: SSB and fishing mortality are strongly influenced by the length composition in the catches: this information create an upward rescaling of the assessment. Recruitment, on the other hand, remains fairly stable.

### 7.7 Management Options

### 7.7.1 MSY approach

In 2017 stochastic equilibrium analyses were carried out using the eqSim software (WKWIDE 2017) to provide an estimate for Fmsy. Since there is no clear evidence of
stock recruitment relationship for western horse mackerel, the stock has been considered as a type 5 (ICES guidelines), so $B_{\text {lim }}$ was set equal to Bloss. A segmented regression S-R was used excluding the 1982 year class and setting the breakpoint at Blim . Biological parameters (mean weights at age, maturity and natural mortality) and exploitation pattern were as in the last 10 years (2006-2015) of the stock assessment. Assessment and advice error for the estimation of the MSY reference points were fixed as the default value used during the WKMSYREF4 ( $\mathrm{F}_{\mathrm{cv}}=0.212 ; \mathrm{F}_{\mathrm{phi}}=0.433$, estimated as the median of 5 stocks). The FMSY was estimated equal to 0.1079 .

During WGWIDE 2017 further investigations on these reference points and the methodology used were carried out: these are summarised in the Working Document attached to WGWIDE report (ICES, 2017a).

### 7.7.2 Management plans and evaluations

An overview of earlier management plans and management plans evaluations was presented at WGWIDE 2017. To date, no agreed management plan is available for this stock despite several attempts to develop such management plans.

New work on the development of a potential Harvest Control Rule (HCR) for Western horse mackerel has been initiated by the Pelagic Advisory in 2018. The PELAC requested Landmark Fisheries Research (Canada) to develop a proof-of-concept of a Management Strategy Evaluation (MSE) testing different types of HCRs. Previously, Landmark Fisheries Research has done similar exercises for e.g. Sablefish in British Columbia.

The approach presented by Landmark Fisheries Research was based on a full-feedback MSE with an embedded stock assessment model included. The approach explicitly recognizes both biomass and fishery objectives. Simulated outcomes under alternative rebuilding plans defined by alternative harvest control rules can be used to examine potential trade-offs among stock rebuilding and fishery performance objectives in both the short and long-term. As expected, risks of the stock being below Blim are highest in the short term; however, long term performance clearly demonstrates the precautionary aspects of the simulated rebuilding plans. In particular, all harvest control rules lead to stock growth in the long term, but with different outcomes in terms of yield, yield variability, and probably of fishery closure. Although some rules led to more rapid stock growth, they did so at considerably higher cost to the fishery than other rules. So far no uncertainty in the initial conditions have been included. Nevertheless, these results suggest that a full MSE could be used to identify management procedures that provide acceptable trade-offs between fishing and spawning biomass conservation of Western horse mackerel.

### 7.8 Management considerations

The 2001 year class has now entered the plus group and there are no detectable strong year classes entering the fishery, even though a higher amount of age 1 and 2 fish has recently been observed in the catches.

With the inclusion of the 2017 length frequency distribution -which, together with the LFD from 2015-2016, shows a change in selectivity toward smaller fish - the SSB is rescaled to higher levels. This rescaling affects the perception of the stock relative to the reference points estimated at the benchmark: even though the 2017 SSB estimate is the lowest point of the time-series, it is now higher than the Blim reference point and close to $B_{\text {trigger. }}$. This, together with the decrease in fishing mortality in the last 3 years, implies an advice which disregard the overall decreasing trend of the population.

The way the current reference points have been estimated is strongly driven by the assumptions on the stock recruitment relationships: the stock was considered as a type 5, therefore Blim was set equal to Bloss, i.e. the SSB in 2015. In general, the last year in a stock assessment model is estimated with a higher uncertainty and it is more unstable due to the lower amount of information compared to earlier years in the time-series: in a stock that has always suffered of retrospective problems, such as the case of Western Horse mackerel, this issue is even more relevant. The use of the last year for the reference point estimation is therefore not advisable and can cause biases in the advice.

The TAC has only been given for parts of the distribution and fishing areas (EU waters). The Working Group advises that the TAC should apply to all areas where western horse mackerel are caught. Note that subarea 8.c is now included in the Western stock distribution area. If (as planned) the management area limits are revised, measures should be taken to ensure that misreporting of juvenile catch taken in subareas 7.e,h and 7.d (the latter then belonging to the North Sea stock management area) is effectively hindered. The mismatch between TAC and fishing areas and the fact that the TAC is only applied to EU waters has resulted in the catch prior to 2007 exceeding those advised by ICES.

The management plan proposed by the Pelagic RAC in 2007 was evaluated by ICES and considered to be precautionary in the short term. This plan makes use of the information available in the egg production surveys, and bases triennial TACs on the slope of the three previous egg production estimates. The rule proposed by the plan was used to set the TAC for 2008-2010 at 180kt. Using the finalised egg survey time-series the catch advice for 2014-2016 is 137534 t . It should be noted that the management plan assumes that all catches are taken against the TAC and, should the management and assessment areas be combined in the future, the TAC as set by the EU will not cover all fisheries. Following an evaluation in 2013, ICES considered this management plan is not precautionary.

### 7.9 Ecosystem considerations

Knowledge about the distribution of the western horse mackerel stock is mostly gained from the egg surveys and the seasonal changes in the fishery. Based on these observations it is not possible to infer a similar changing trend in the distribution of western horse mackerel as for NEA mackerel. However, from catch data it appears that the stock is concentrating in the southern areas and it is mostly characterized by small individuals.

### 7.10 Regulations and their effects

There are no horse mackerel management agreements between EU and non EU countries. The TAC set by EU therefore only apply to EU waters and the EU fleet in international waters. The minimum landing size of horse mackerel by the EU fleet is 15 cm ( $10 \%$ undersized allowed in the catches).

The stock allocations were changed in 2005 following the results of the HOMSIR project (Abaunza et al. 2003) and 8.c now belongs to the western stock. Landings from 7.d are now allocated to the North Sea horse mackerel. A research project is currently underway in the Netherlands and Ireland, to review the stock separation between the Western stock and the North Sea stock in the Channel area (see North Sea horse mackerel section in the report). Results are expected to be available in the first half of 2019.

In Norwegian waters there is no quota for horse mackerel but existing regulations on bycatch proportions as well as a general discard prohibition (for all species) apply to horse mackerel.

### 7.11 Changes in fishing technology and fishing patterns

The description of the fishery is given in Sections 5.1 and 7.2.1 and no large changes in fishing areas or patterns have taken place. However, there has been a gradual shift from an industrial fishery for meal and oil towards a human consumption fishery.

### 7.12 Changes in the environment

Migrations are closely associated with the slope current, and horse mackerel migrations are known to be modulated by temperature. Continued warming of the slope current is likely to affect the timing and spatial extent of this migration.

Since the strong 1982 year class of the western stock started to appear in the North Sea in 1987 a good correspondence between the modelled influx of Atlantic water to the North Sea in the first quarter and the horse mackerel catches taken by Norwegian purse-seiners in the Norwegian EEZ (NEZ) later (October-November) the same year (Iversen et al. 2002, Iversen WD presented in ICES 2007/ACFM:31) has been noted in most years.

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### 7.14 Tables

Table 7.1.1.1. Western horse mackerel. Catches (t) in Subarea 2. (Data as submitted by Working Group members).

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - | - | - | - | 39 |
| France | - | - | - | - | 1 | 1 | - ${ }^{2}$ | -2 |
| Germany, Fed.Rep | - | + | - | - | - | - | - | - |
| Norway | - | - | - | 412 | 22 | 78 | 214 | 3,272 |
| USSR | - | - | - | - | - | - | - | - |
| Total | - | + | - | 412 | 23 | 79 | 214 | 3,311 |
|  | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| Faroe Islands | - | - | 9643 | 1,115 | 9,157 ${ }^{3}$ | 1,068 | - | 950 |
| Denmark | - | - | - | - | - | - | - | 200 |
| France | -2 | - | - | - | - | - | 55 | - |
| Germany, Fed. Rep. | 64 | 12 | + | - | - | - | - | - |
| Norway | 6,285 | 4,770 | 9,135 | 3,200 | 4,300 | 2,100 | 4 | 11,300 |
| USSR / Russia (1992-) | 469 | 27 | 1,298 | 172 | - | - | 700 | 1,633 |
| UK (England + Wales) | - | - | 17 |  | - | - | - | - |
| Total | 6,818 | 4,809 | 11,414 | 4,487 | 13,457 | 3,168 | 759 | 14,083 |
|  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| Faroe Islands | 1,598 | 7993 | $188{ }^{3}$ | $132{ }^{3}$ |  | - | - | - |
| Denmark | - | - | 1,755 ${ }^{3}$ | - |  | - | - | - |
| France | - | - | - | - |  | - | - | - |
| Germany | - | - | - | - |  | - | - | - |
| Norway | 887 | 1,170 | 234 | 2,304 | 841 | 44 | 1,321 | 22 |
| Russia | 881 | 554 | 345 | 121 | 78 | 16 | 3 | 2 |
| UK (England + Wales) | - | - | - | - | - | - | - | - |
| Estonia | - | 78 | 22 | - | - | - | - | - |
| Total | 3,366 | 2,601 | 2,544 | 2557 | 919 | 60 | 1,324 | 24 |
|  | 2004 | 2005 |  | 2006 | 2007 | 2008 | 2009 | 2010 |
| Faroe Islands | - | - | - | 3 | - | - | - | 222 |
| Denmark | - | - | - | - | - | - | - | - |
| France | - | - | - | - | - | - | - | - |
| Germany | - | - | - | - |  | - | - | - |
| Ireland | - | - | - | - |  | - | - | - |
| Norway | 42 | 176 |  | 27 | - | 572 | 1,847 | 1,364 |
| Russia | - | - | - | - | - | - | - | - |
| UK (England + Wales) | - | - | - | - | - | - | - | - |
| Estonia | - | - | - | - | - | - | - | - |
| Total | 42 | 176 |  | 27 | 0 | 572 | 1,847 | 1,586 |
| ${ }^{2}$ Included in Subarea 4. <br> ${ }^{3}$ Includes catches in Div. ${ }^{4}$ Taken in Div. 5.b |  |  |  |  |  |  |  |  |

Table 7.1.1.1 cont. Western horse mackerel. Catches ( t ) in Subarea 2. (Data as submitted by Working Group members).

|  | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | $2017{ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 224 | - | - | - | - |  |  |
| Denmark | - | - | - | - | - |  |  |
| France | - | + | - | - | - |  |  |
| Germany | - | - | - | - | - |  |  |
| Ireland | - | - | - | - | - |  |  |
| Netherlands | 1 | - |  | 107 | - |  |  |
| Norway | 298 | 66 | 30 | 302 | 10 | 45 | 5 |
| Russia | - | - |  | - | - |  |  |
| UK (England + Wales) | - | - |  | - | - |  |  |
| Estonia | - | - |  | - | - |  |  |
| Total | 523 | 66 | 30 | 409 | 10 | 45 | 5 |
| ${ }^{1}$ Preliminary <br> ${ }^{2}$ Included in 4. <br> ${ }^{3}$ Includes catches in Div <br> ${ }^{4}$ Taken in Div. 5.b. |  |  |  |  |  |  |  |

Table 7.1.1.2. Western horse mackerel. Catches ( t ) in North Sea Subarea 4 and Skagerrak Division 3.a by country. (Data submitted by Working Group members). Catches partly concern the North Sea horse mackerel.

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 8 | $34 \quad 7$ | 7 | 55 | 20 | 13 | 13 | 9 | 10 |
| Denmark | 199 | 3,576 1, | 1,612 | 1,590 | 23,730 | 22,495 | 18,652 | 7,290 | 20,323 |
| Faroe Islands | 260 | - - | - - | - - | - - | - - | - | - | - |
| France | 292 | 4215 | 567 | 366 | 827 | 298 | $231{ }^{2}$ | 1891 | $784{ }^{1}$ |
| Germany, Fed.Rep. | + | 1393 | 30 | 52 | + + | + - | - | 3 | 153 |
| Ireland | 1,161 | 412 | - - | - - | - - | - - | - | - | - |
| Netherlands | 101 | 355 | 559 | 2,029 ${ }^{2}$ | 824 | $160^{2}$ | $600^{2}$ | $850^{3}$ | 1,060 ${ }^{3}$ |
| Norway ${ }^{2}$ | 119 | 2,292 7 | 7 | 322 | 2 | 203 | 776 | $11,728^{3}$ | $34,425^{3}$ |
| Poland | - - | - - | - 2 | 29 | 94 | - - | - | - | - |
| Sweden | - | - - | - - | - - | - - | - 2 | 2 | - | - |
| UK (Engl. + Wales) | 11 | 156 | 6 | 4 - | - 7 | 71 | 3 | 339 | 373 |
| UK (Scotland) | - | - - | - - | - 3 | 3 | 9985 | 531 | 487 | 5,749 |
| USSR | - | - - | - - | - 48 | 489 | - - | - | - | - |
| Total | 2,151 | 7,253 | 2,788 | 4,420 | 25,987 | 24,238 | 20,808 | 20,895 | 62,877 |
| COUNTRY | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| Belgium | 10 | 13 | - | + | 74 | 57 | 51 | 28 | - |
| Denmark | 23,329 | 20,605 | 6,982 | 7,755 | 6,120 | 3,921 | 2,432 | 1,433 | 976 |
| Estonia | - | - | - | 293 | - |  | 17 | - | - |
| Faroe Islands | - | 942 | 340 | - | 360 | 275 | - | - | 296 |
| France | 248 | 220 | 174 | 162 | 302 |  | - | - | - |
| Germany, Fed.Rep. | 506 | 2,4694 | 5,995 | 2,801 | 1,570 | 1,014 | 1,600 | 7 | 37 |
| Ireland | - | 687 | 2,657 | 2,600 | 4,086 | 415 | 220 | 1,100 | 8,152 |
| Netherlands | 14,172 | 1,970 | 3,852 | 3,000 | 2,470 | 1,329 | 5,285 | 6,205 | 52 |
| Norway | 84,161 | 117,903 | 50,000 | 96,000 | 126,800 | 94,000 | 84,747 | 14,639 | 43,888 |
| Poland | - | - | - | - | - | - | - | - | - |
| Sweden | - | 102 | 953 | 800 | 697 | 2,087 | - | 95 | 1761 |
| UK (Engl. + Wales) | 10 | 10 | 132 | 4 | 115 | 389 | 478 | 40 | 10 |
| UK (N. Ireland) | - | - | 350 | - | - |  | - | - | - |
| UK (Scotland) | 2,093 | 458 | 7,309 | 996 | 1,059 | 7,582 | 3,650 | 2,442 | 10,511 |
| USSR / Russia (1992- | - | - | - |  |  |  |  |  |  |
| ) | 12,4823 | -3173 | -7503 | -2785 | -3,270 | 1,511 | -28 | 136 | -31,615 |
| Unallocated+discards |  |  |  |  |  |  |  |  |  |
| Total | 112,047 | 145,062 | 77,904 | 114,133 | 140,383 | 112,580 | 98,452 | 26,125 | 34,068 |


| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | 19 | 21 |  |  |  |  |  |  |  |
| Denmark | 2048 | 2026 | 7 | 98 | 53 | 841 | 48 | 216 | 60 |
| Estonia |  | - |  |  |  |  |  |  |  |
| Faroe Islands | 28 | 908 | 24 | 0 | 671 | 5 | 76 | 35 | 0 |
| France | 379 | 60 | 49 |  |  | 255 |  | 1 |  |
| Germany | 4620 | 4072 | 0 | 0 | 4 | 534 | 0 | 44 | 1 |
| Ireland | - | 404 | 32 | 332 | 11 | 93 | 378 |  |  |
| Lithuania |  |  |  |  |  |  |  |  |  |
| Netherlands | 4548 | 3285 | 10 | 1 | 0 | 36 | 0 | 0 | 0 |


| Norway | 13129 | 44344 | 1141 | 7912 | 34843 | 20349 | 10687 | 24733 | 27087 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Russia | - | - | 2 |  |  |  |  |  |  |
| Sweden | 1761 | 1957 | 1009 | 68 | 561 | 1002 | 567 | 216 | 0 |
| UK (Engl. + Wales) | 1 | 12 |  |  |  |  | 0 |  |  |
| UK (Scotland) | 3041 | 1658 | 3054 | 3161 | 252 | 0 | 0 | 22 | 61 |
| Unallocated+discards | 737 | -325 | 10 | 0 | 0 | -36 | 0 | 0 | 0 |
| Total | 30311 | 58422 | 5338 | 11572 | 36395 | 23079 | 11756 | 25267 | 27210 |

${ }^{1}$ Includes Division 2.a. ${ }^{2}$ Estimated from biological sampling. ${ }^{3}$ Assumed to be misreported. ${ }^{4}$ Includes 13 t from the German Democratic Republic. ${ }^{5}$ Includes a negative unallocated catch of $4,000 \mathrm{t}$.

Table 7.1.1.2 cont. Western horse mackerel. Catches (t) in North Sea Subarea 4 and Skagerrak Division 3.a by country. (Data submitted by Working Group members). Catches partly concern the North Sea horse mackerel.

| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Denmark | 74 | 2 | 207 | 61 | 19 | 9 | 0 | 23 |
| Faroe Islands | 3 | 55 | 0 | 8 | 0 | 0 | 0 | 53 |
| France |  | 1 |  |  | 268 |  | 17 |  |
| Germany, Fed.Rep. | 6 | 93 | 0 | 4 | 0 | 0 | 20 | 0 |
| Ireland | 651 | 298 | 342 | 14 | 755 | 25 | 7 |  |
| Netherlands |  |  |  |  |  |  | 0 | 0 |
| Lithuania | 22 | 0 | 7 | 339 | 81 | 92 | 0 | 310 |
| Norway | 4180 | 11631 | 57890 | 10556 | 13409 | 3183 | 6566 | 14051 |
| Sweden | 76 | 9 | 258 | 2 | 90 | 0 | 1 | 0 |
| UK (Engl. + Wales) | 31 |  |  |  |  |  | 16 | 203 |
| UK (Scotland) | 7 | 20 | 51 | 546 | 101 | 12 | 102 | 11 |
| Unallocated +discards | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 30 |
| Total | 5050 | 12110 | 58755 | 11531 | 14723 | 3320 | 6712 | 14699 |
| 1-Preliminary. |  |  |  |  |  |  | 0 |  |

Table 7.1.1.2 cont. Western horse mackerel. Catches ( $t$ ) in North Sea Subarea 4 and Skagerrak Division 3.a by country. (Data submitted by Working Group members). Catches partly concern the North Sea horse mackerel.

| Country | 2015 | 2016 | $2017^{1}$ |
| :--- | ---: | ---: | ---: |
| Denmark | 37 | 7 | 21 |
| Faroe Islands | 0 | 0 | 67 |
| France | 12 | 4 | 1 |
| Germany, | 6 | 28 | 1 |
| Fed.Rep. | 8 |  |  |
| Ireland | 12 | 130 |  |
| Netherlands | 0887 | 8765 | 9880 |
| Lithuania | 10 | 0 | 41 |
| Norway | 134 | 13 | 4 |
| Sweden | 36 | 14 |  |
| UK (Engl. + | 32 | 97 | 87 |
| Wales) | 9175 | 9057 | 10117 |
| UK (Scotland) |  |  |  |

Table 7.1.1.3 Western horse mackerel. Catches ( $t$ ) in Subarea 6 by country. (Data submitted by Working Group members).

| Country | 1980 | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Denmark | 734 | 341 | 2,785 | 7 | - | - | - | 769 | 1,655 |
| Faroe Islands | - | - | 1,248 | - | - | 4,014 | 1,992 | $4,450^{2}$ | $4,000^{2}$ |
| France | 45 | 454 | 4 | 10 | 14 | 13 | 12 | 20 | 10 |
| Germany, Fed. Rep. | 5,550 | 10,212 | 2,113 | 4,146 | 130 | 191 | 354 | 174 | 615 |
| Ireland | - | - | - | 15,086 | 13,858 | 27,102 | 28,125 | 29,743 | 27,872 |
| Netherlands | 2,385 | 100 | 50 | 94 | 17,500 | 18,450 | 3,450 | 5,750 | 3,340 |
| Norway | - | 5 | - | - | - |  | 83 | 75 | 41 |
| Spain | - | - | - | - | - |  | -1 | -1 | -1 |
| UK (Engl. + Wales) | 9 | 5 | + | 38 | + | 996 | 198 | 404 | 475 |
| UK (N. Ireland) |  |  |  |  |  |  |  | - | - |

${ }^{1}$ Included in Subarea 7. ${ }^{2}$ Includes Divisions 3.a, 4.a, b and 6.b. ${ }^{3}$ Includes a negative unallocated catch of 7000 t.

Table 7.1.1.3. cont. Western horse mackerel. Catches ( $t$ ) in Subarea 6 by country. (Data submitted by Working Group members).

| Country | 2007 | 2008 | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 5}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Denmark |  |  |  |  | 58 | 1,131 | 433 | 856 | 3,045 |
| Faroe Islands |  | 573 |  | 66 |  |  |  |  |  |
| France |  | 73 |  |  | 246 |  |  | 195 | 65 |
| Germany | 1,835 | 5,097 | 635 | 773 | 6,508 | 671 | 8,616 | 4,194 | 1,980 |
| Ireland | 20,010 | 18,751 | 16,596 | 19,985 | 23,556 | 29,282 | 19,979 | 15,745 | 10,894 |
| Lithuania | 80 | 641 |  |  |  |  |  |  |  |
| Netherlands | 2,177 | 3,904 | 2,332 | 1,684 | 6,353 | 12,653 | 11,078 | 8,580 | 6,211 |
| Norway | 2 | 20 | 27 | 18 | 48 | 2 |  |  |  |
| Spain | 0 |  |  |  |  |  |  |  |  |
| UK (Engl. + Wales) | 332 |  |  | 463 |  |  | 451 | 18 | 58 |
| UK (N.Ireland) |  |  |  | 59 | 198 |  | 2,325 | 1,579 | 1,204 |
| UK (Scotland) | 38 | 588 | 243 | 89 | 2,528 | 1,231 | 385 | 1,277 | 696 |
| Unallocated+disc. | 0 | 0 | 0 | 0 | 230 | 2 | - | 123 |  |
| Total | 24,474 | 29,648 | 19,833 | 23,136 | 39,726 | 44,973 | 43,266 | 32,567 | 24,153 |

${ }^{1}$ Preliminary.

Table 7.1.1.3. cont. Western horse mackerel. Catches ( $\mathbf{t}$ ) in Subarea 6 by country. (Data submitted by Working Group members).

| Country | 2016 | $2017^{1}$ |
| :--- | ---: | ---: |
| Denmark | 3,462 |  |
| Faroe Islands | 113 |  |
| France | 23 | 1,025 |
| Germany | 4,069 | 2,884 |
| Ireland | 15,381 | 15,123 |
| Lithuania | 2,510 |  |
| Netherlands | 9,246 | 5,497 |
| Norway |  |  |
| Spain |  |  |
| UK (Engl. + Wales) |  |  |
| UK (N.Ireland) | 0 |  |
| UK (Scotland) | 956 |  |
| Unallocated+disc. |  | 116 |
| Total |  |  |

${ }^{1}$ Preliminary.

Table 7.1.1.4. Western horse mackerel. Catches (t) in Subarea 7 by country. (Data submitted by the Working Group members).


Table 7.1.1.4. cont. Western horse mackerel. Catches ( $t$ ) in Subarea 7 by country. (Data submitted by the Working Group members).

| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 475 | 212 |  | - | - | - | 0 |  |  |
| Belgium |  |  |  | 19 | 2 |  | 14 |  |  |
| Denmark | 4856 | 1970 | 2710 | 5247 | 5831 | 2281 | 6373 | 5066 | 1474 |
| France | 2007 | 9703 |  | 260 | 7431 | 579 | 744 | 940 | 1552 |
| Germany | 3943 | 5693 | 14205 | 16847 | 14545 | 16391 | 15781 | 12948 | 7382 |
| Ireland | 8039 | 16282 | 23816 | 24491 | 14154 | 15893 | 15805 | 16922 | 10751 |
| Lithuania | 5387 | 4907 |  |  |  | - | 0 |  |  |
| Netherlands | 32654 | 28077 | 23263 | 65865 | 49207 | 53644 | 41562 | 15529 | 18100 |
| Norway | - | - | - | 40 |  | - | 0 |  |  |
| Spain | 11 | 11 | 6 | 3 |  | 10 | 0 |  |  |
| UK (Engl. + Wales) | 5119 | 3245 | 6257 | 12139 | 11688 | 12122 | 3388 | 4576 | 1798 |
| UK (Scotland) |  | 469 | 1119 | 1713 | 299 | 91 | 17 | 101 | 6 |
| Unallocated+discards | 6012 | -4624 | -10891 | 6511 | 1 | 3038 | 4399 | 974 | 1929 |
| Total | 68504 | 65946 | 60487 | 133136 | 103157 | 104049 | 88083 | 57055 | 42992 |
| Country | 2016 | $2017{ }^{1}$ |  |  |  |  |  |  |  |
| Faroe Islands |  |  |  |  |  |  |  |  |  |
| Belgium |  |  |  |  |  |  |  |  |  |
| Denmark | 314 | 1057 |  |  |  |  |  |  |  |
| France | 551 | 595 |  |  |  |  |  |  |  |
| Germany | 7313 | 4077 |  |  |  |  |  |  |  |
| Ireland | 12193 | 7857 |  |  |  |  |  |  |  |
| Lithuania | 86 |  |  |  |  |  |  |  |  |
| Netherlands | 14415 | 8445 |  |  |  |  |  |  |  |
| Norway |  |  |  |  |  |  |  |  |  |
| Spain | 0 |  |  |  |  |  |  |  |  |
| UK (Engl. + Wales) | 820 | 478 |  |  |  |  |  |  |  |
| UK (Scotland) |  |  |  |  |  |  |  |  |  |
| Unallocated+discards | 1692 | 830 |  |  |  |  |  |  |  |
| Total | 37384 | 23340 |  |  |  |  |  |  |  |

${ }^{1}$ Preliminary. ${ }^{2}$ French catches landed in the Netherlands

Table 7.1.1.5. Western horse mackerel. Catches ( $\mathbf{t}$ ) in Subarea 8 by country. (Data submitted by Working Group members).

| Country | 1980 | 1981 | 1982 | 1983 | 1984 |  | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - | - - | - | - | 446 | 3,283 | 2,793 |
| France | 3,361 | 3,711 | 3.073 | 2,643 | 2,489 |  | 4,305 | 3,534 | 3,983 | 4,502 |
| Netherlands | - | - | - | - | -2 | -2 | -2 | -2 | -2 | - |
| Spain | 34,134 | 36,362 | 19,610 | 25,580 | - 23,119 |  | 23,292 | 40,334 | 30,098 | 26,629 |
| UK (Engl.+Wales) | - | + | 1 | - | 1 | 1 | 143 | 392 | 339 | 253 |
| USSR | - | - | - | - | 20 | 20 | - | 656 | - | - |
| Total | 37,495 | 40,073 | 22,684 | 28,223 | 25,629 |  | 27,740 | 45,362 | 37,703 | 34,177 |
| Country | 1989 | 1990 | 1991 | 1992 | 1993 |  | 1994 | 1995 | 1996 | 1997 |
| Denmark | 6,729 | 5,726 | 1,349 | 5,778 | 1,955 |  | - | 340 | 140 | 729 |
| France | 4,719 | 5,082 | 6,164 | 6,220 | -4,010 |  | 28 | - | 7 | 8,564 |
| Germany, Fed. Rep. | - | - | 80 | 62 |  | - |  | - | - | - |
| Netherlands | - | 6,000 | 12,437 | 9,339 | 19,000 |  | 7,272 | - | 14,187 | - |
| Spain | 27,170 | 25,182 | 23,733 | 27,688 | 27,921 |  | 25,409 | 28,349 | 29,428 | 31,082 |
| UK (Engl.+Wales) | 68 | 6 | 70 | 88 | -123 | 23 | 753 | 20 | 924 | 430 |
| Unallocated+discards | - | 1,500 | 2,563 | 5,011 | 1700 | 00 | 2,038 | - | 3,583 | -2,944 |
| Total | 38,686 | 43,496 | 46,396 | 54,186 | 53,709 |  | 35,500 | 28,709 | 48,269 | 37,861 |
| Country | 1998 | 1999 | 2000 | 2001 | 2002 |  | 2003 | 2004 | 2005 | 2006 |
| Denmark | 1,728 | 4,769 | 2,584 | 582 |  |  |  |  |  | 1,513 |
| France | 1,844 | 74 | 7 | 5,316 | 13,676 |  | 4,908 | 2,161 | 3,540 | 3,944 |
| Germany | 3,268 | 3,197 | 3,760 | 3,645 | 2,293 |  | 504 | 72 | 4,776 | 3,326 |
| Ireland | - | - | 6,485 | 1,483 | -704 | 04 | 1,314 | 1,882 | 1,808 | 158 |
| Lithuania | - | - |  |  |  |  |  |  |  | 401 |
| Netherlands | 8,123 | 13,821 | 11,769 | 35,106 | 12,538 |  | 6,620 | 1,047 | 6,372 | 6,073 |
| Spain | 23,599 | 24,461 | 24,154 | 23,531 | 24,752 |  | 24,598 | 16,245 | 16,624 | 13,874 |
| UK (Engl. + Wales) | 9 | 28 | 121 | 1,092 | - 1,578 |  | 982 | 516 | 838 | 821 |
| UK (Scotland) | - | - | 249 |  |  |  |  |  |  |  |
| Unallocated+discards | 1,884 | -8658 | 5,093 | 4,365 | 1,705 |  | 2,785 | 2,202 | 7,302 | 4,013 |
| Total | 40,455 | 37,692 | 54,222 | 75,120 | 57,246 |  | 41,711 | 24,125 | 41,260 | 34,122 |
| Country 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 32014 | 2015 | 2016 | $2017{ }^{1}$ |
| Denmark 2,687 | 3,289 | 3,109 | 632 | 200 | 581 |  | 4 |  |  | 1 |
| France 10,741 | 2,848 |  |  | 326 | 1,218 | 2,849 | 2,277 | 1,618 | 2,219 | 2,303 |
| Germany | 918 | 281 | 64 | 61 |  | 417 | 719 | 949 | 4 | 210 |
| Ireland 694 |  |  |  |  | 39 |  |  | 0 | 32 | 580 |
| Netherlands 211 | 6,269 | 1,848 | 98 | 49 | 7 | 1,057 | 57526 | 6635 | 1 | 313 |
| Spain 14,265 | 19,840 | 21,071 | 38,742 | 34,581 13 | 13,502 2 | 22,542 | 19,443 | 13,072 | 14,235 | 14,901 |
| UK (Engl. + Wales) | 120 | 224 | 112 | 28 |  | 104 | 4 35 | 572 | 9 |  |
| Unallocated+discards | 67 | 913 | 7,412 | 417 | 431 | 2,055 | 55182 | 29,314 | 6,643 | 2,907 |
| Total 28,598 | 33,352 | 27,447 | 47,060 | 35,662 1 | 15,777 2 | 29,039 | 22,483 | 24,760 | 23,143 | 21,213 |

${ }^{1}$ Preliminary. ${ }^{2}$ Included in Subarea 7. ${ }^{3}$ French catches landed in the Netherlands

Table 7.2.1.1. Western horse mackerel. The time-series of Total Annual Egg Production (TAEP) estimates ( $\mathbf{1 0}^{12}$ eggs).

| Year | TAEP |
| :--- | :--- |
| 1992 | 2094 |
| 1995 | 1344 |
| 1998 | 1242 |
| 2001 | 864 |
| 2004 | 884 |
| 2007 | 1486 |
| 2010 | 1033 |
| 2016 | 366 |

Table 7.2.4.1. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2017

| Q1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.5.b.1.b | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | total |
| 0 |  |  |  |  |  |  |  |  | 0 |  |  |  |  | 0 |
| 1 |  |  |  |  |  |  |  |  | 0 |  |  |  |  | 0 |
| 2 | 0 |  | 130 | 0 |  |  | 79 | 333 | 0 | 1 | 1403 | 15 | 12 | 1973 |
| 3 | 7 |  | 32049 | 80 | 708 |  | 872 | 3662 | 0 | 14 | 13463 | 165 | 131 | 51151 |
| 4 | 1 |  | 6835 | 17 | 1495 |  | 369 | 166 | 0 | 4 | 4130 | 70 | 55 | 13143 |
| 5 | 2 |  | 10081 | 25 | 13541 | 166 | 1174 |  | 0 | 2 | 2196 | 223 | 177 | 27588 |
| 6 | 0 |  | 2523 | 6 | 3269 | 76 | 292 |  | 0 | 1 | 709 | 55 | 44 | 6975 |
| 7 | 0 |  | 1635 | 4 | 3081 |  | 258 |  |  | 0 | 370 | 49 | 39 | 5436 |
| 8 | 1 |  | 6059 | 15 | 3920 | 277 | 348 |  | 0 | 1 | 578 | 68 | 54 | 11320 |
| 9 | 2 |  | 16445 | 41 | 16373 | 689 | 1310 |  | 0 | 0 | 496 | 251 | 199 | 35806 |
| 10 | 0 |  | 2246 | 6 | 1930 | 118 | 154 |  | 0 |  |  | 29 | 23 | 4506 |
| 11 | 0 |  | 916 | 2 | 872 | 38 | 81 |  | 0 | 0 | 84 | 14 | 11 | 2019 |
| 12 | 0 |  | 1018 | 3 | 398 | 83 | 44 |  | 0 | 0 | 74 | 8 | 7 | 1635 |
| 13 | 0 |  | 1391 | 3 | 1067 | 194 | 99 |  | 0 | 0 | 213 | 20 | 16 | 3004 |
| 14 | 0 |  | 2028 | 5 | 389 | 76 | 38 |  | 0 |  |  | 7 | 6 | 2550 |
| 15 | 1 |  | 10360 | 26 | 4813 | 273 | 414 |  | 0 | 0 | 70 | 77 | 61 | 16096 |
| sum | 16 |  | 93717 | 234 | 51857 | 1991 | 5532 | 4161 | 0 | 24 | 23785 | 1051 | 833 | 183203 |

Table 7.2.4.1 cont. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2017

| Q2 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.5.b.1.b | 27.6.a | 27.7.b | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k. 2 | total |
| 0 |  |  |  |  | 267 |  |  |  |  |  |  | 267 |
| 1 |  |  |  |  | 335 |  |  |  |  |  |  | 335 |
| 2 |  | 1 |  | 1 | 229 |  | 21 | 29 | 371 | 321 | 1 | 974 |
| 3 |  | 160 |  | 11 | 1211 |  | 204 | 284 | 3610 | 3123 | 5 | 8609 |
| 4 |  | 24 |  | 5 | 26 |  | 57 | 79 | 1007 | 872 | 1 | 2071 |
| 5 |  | 36 |  | 15 | 38 |  | 30 | 41 | 527 | 456 | 1 | 1144 |
| 6 |  | 8 |  | 4 |  |  | 10 | 13 | 170 | 147 | 0 | 353 |
| 7 |  | 17 | 42 | 3 |  |  | 5 | 7 | 90 | 77 | 0 | 242 |
| 8 |  | 32 | 42 | 5 |  |  | 8 | 11 | 140 | 121 | 0 | 358 |
| 9 |  | 90 | 124 | 17 |  |  | 7 | 9 | 120 | 104 | 0 | 471 |
| 10 |  | 32 | 82 | 2 |  |  |  | 0 | 2 | 1 | 0 | 119 |
| 11 |  | 26 | 80 | 1 |  |  | 1 | 2 | 21 | 18 | 0 | 150 |
| 12 |  | 27 | 82 | 1 |  |  | 1 | 1 | 18 | 16 | 0 | 146 |
| 13 |  | 16 | 40 | 1 |  |  | 3 | 4 | 51 | 44 | 0 | 160 |
| 14 |  | 54 | 164 | 0 |  |  |  | 0 | 0 | 0 | 0 | 219 |
| 15 |  | 422 | 1354 | 5 |  |  | 1 | 1 | 17 | 15 | 0 | 1816 |
| sum |  | 945 | 2011 | 72 | 2107 |  | 347 | 484 | 6145 | 5316 | 9 | 17435 |

Table 7.2.4.1 cont. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2017

| Q3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.3.a | 27.4.a | 27.5.b.1.b | 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.k. 2 | total |
| 0 |  |  |  |  |  |  |  | 692 | 39 | 49 |  |  |  | 779 |
| 1 | 533 |  |  |  |  |  |  | 867 | 48 | 61 |  |  |  | 1509 |
| 2 | 486 |  |  | 1 |  | 0 | 0 | 593 | 28 | 35 |  |  |  | 1143 |
| 3 | 1485 | 335 |  | 228 | 362 | 3 | 4 | 3133 | 116 | 147 | 67 | 11 | 1 | 5893 |
| 4 | 291 | 1617 |  | 28 | 89 | 1 | 2 | 67 | 1 | 1 | 17 | 3 | 0 | 2116 |
| 5 | 214 | 1733 |  | 45 | 20 | 4 | 6 | 99 | 6 | 7 | 4 | 1 | 0 | 2138 |
| 6 | 90 | 168 |  | 9 | 14 | 1 | 1 |  | 2 |  | 3 | 0 | 0 | 288 |
| 7 | 90 | 655 |  | 7 |  | 1 | 1 |  |  |  |  |  |  | 755 |
| 8 | 163 | 2340 |  | 24 | 25 | 1 | 2 |  | 2 | 1 | 5 | 1 | 0 | 2563 |
| 9 | 247 | 4845 |  | 63 | 69 | 4 | 7 |  | 15 | 1 | 13 | 2 | 0 | 5266 |
| 10 | 116 | 2359 |  | 11 | 10 | 1 | 1 |  | 5 | 0 | 2 | 0 | 0 | 2504 |
| 11 | 40 | 590 |  | 4 |  | 0 | 0 |  | 7 | 0 |  |  |  | 641 |
| 12 | 26 | 550 |  | 4 | 10 | 0 | 0 |  | 3 | 0 | 2 | 0 | 0 | 596 |
| 13 | 53 | 1143 |  | 5 | 5 | 0 | 0 |  | 4 | 0 | 1 | 0 | 0 | 1212 |
| 14 | 71 | 1401 |  | 9 |  | 0 | 0 |  | 3 |  |  |  |  | 1484 |
| 15 | 324 | 6814 |  | 41 | 40 | 1 | 2 |  | 15 | 0 | 7 | 1 | 0 | 7247 |
| sum | 4230 | 24550 |  | 479 | 645 | 19 | 27 | 5451 | 291 | 302 | 120 | 20 | 2 | 36136 |

Table 7.2.4.1 cont. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2017

| Q4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.3.a | 27.4.a | 27.5.b.1.b | 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.k | 27.7.k. 2 | total |
| 0 |  |  |  |  |  |  |  |  | 1400 | 6 |  |  |  |  |  | 1406 |
| 1 |  | 3 |  |  |  |  |  |  | 1755 | 8 |  |  |  |  |  | 1765 |
| 2 | 0 | 3 | 20 |  | 94 |  | 6 | 8 | 978 | 4 |  | 15 |  | 0 | 0 | 1129 |
| 3 | 5 | 8 | 7379 |  | 35030 | 802 | 71 | 89 | 3894 | 18 |  | 173 |  | 0 | 4 | 47472 |
| 4 | 1 | 2 | 528 |  | 2521 | 531 | 30 | 38 | 24 | 0 |  |  |  | 0 | 2 | 3676 |
| 5 | 1 | 1 | 917 |  | 4370 | 990 | 95 | 120 | 200 | 1 | 14 |  | 23 | 0 | 5 | 6739 |
| 6 | 0 | 1 | 126 |  | 603 | 122 | 24 | 30 |  | 0 | 32 |  |  | 0 | 1 | 938 |
| 7 | 0 | I | 111 |  | 532 | 199 | 21 | 26 |  |  |  |  |  | 0 | 1 | 891 |
| 8 | 1 | 1 | 403 |  | 1923 | 473 | 28 | 36 |  | 0 | 32 |  | 39 | 0 | 2 | 2936 |
| 9 | 2 | 1 | 1016 |  | 4855 | 898 | 106 | 134 |  | 2 | 266 |  | 62 | 0 | 6 | 7347 |
| 10 | 0 | 1 | 263 |  | 1253 | 83 | 12 | 16 |  | 1 | 82 |  | 8 | 0 | 1 | 1719 |
| 11 | 0 | 0 | 82 |  | 392 | 50 | 7 | 8 |  | 1 | 122 |  | 8 | 0 | 0 | 670 |
| 12 | 0 | 0 | 81 |  | 385 | 138 | 4 | 4 |  | 0 | 53 |  | 8 | 0 | 0 | 673 |
| 13 | 0 | 0 | 64 |  | 306 | 50 | 8 | 10 |  | 1 | 72 |  | 15 | 0 | 0 | 527 |
| 14 | 0 | 0 | 174 |  | 830 | 128 | 3 | 4 |  | 0 | 53 |  |  | 0 | 0 | 1193 |
| 15 | 1 | 2 | 691 |  | 3300 | 720 | 34 | 42 |  | 2 | 261 |  | 31 | 0 | 2 | 5085 |
| sum | 11 | 24 | 11855 |  | 56392 | 5184 | 449 | 566 | 8251 | 45 | 986 | 188 | 193 | 0 | 23 | 84167 |

Table 7.2.4.1 cont. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2017

| Q1-4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.3.a | 27.4.a | 27.5.b.1.b | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k | 27.7.k. 2 | total |
| 0 |  |  |  |  |  |  |  |  |  | 2360 | 45 | 49 |  |  |  |  |  | 2453 |
| 1 |  | 536 |  |  |  |  |  |  |  | 2957 | 56 | 61 |  |  |  |  |  | 3610 |
| 2 | 0 | 489 | 20 |  | 224 | 0 |  | 7 | 89 | 2133 | 32 | 57 | 1447 | 386 | 333 | 0 | 1 | 5217 |
| 3 | 12 | 1494 | 7715 |  | 66987 | 80 | 1871 | 74 | 977 | 11899 | 134 | 365 | 13988 | 3787 | 3254 | 0 | 10 | 112644 |
| 4 | 2 | 293 | 2145 |  | 9337 | 17 | 2116 | 31 | 413 | 283 | 1 | 62 | 4226 | 1080 | 927 | 0 | 3 | 20936 |
| 5 | 3 | 215 | 2650 |  | 14424 | 25 | 14551 | 266 | 1315 | 337 | 7 | 54 | 2241 | 774 | 632 | 0 | 6 | 37499 |
| 6 | 1 | 91 | 294 |  | 3119 | 6 | 3404 | 101 | 327 |  | 2 | 43 | 725 | 226 | 191 | 0 | 2 | 8531 |
| 7 | 0 | 91 | 767 |  | 2174 | 4 | 3322 | 22 | 289 |  |  | 5 | 377 | 138 | 116 | 0 | 1 | 7307 |
| 8 | 1 | 164 | 2743 |  | 7978 | 15 | 4461 | 306 | 390 |  | 2 | 41 | 593 | 247 | 175 | 0 | 2 | 17117 |
| 9 | 4 | 249 | 5861 |  | 21290 | 41 | 17465 | 800 | 1467 |  | 17 | 274 | 518 | 434 | 302 | 0 | 6 | 48728 |
| 10 | 1 | 117 | 2622 |  | 3516 | 6 | 2105 | 131 | 173 |  | 5 | 82 | 2 | 39 | 24 | 0 | 1 | 8823 |
| 11 | 0 | 40 | 672 |  | 1329 | 2 | 1003 | 45 | 91 |  | 8 | 123 | 86 | 43 | 29 | 0 | 0 | 3471 |
| 12 | 0 | 26 | 631 |  | 1423 | 3 | 629 | 87 | 49 |  | 3 | 54 | 77 | 35 | 22 | 0 | 0 | 3040 |
| 13 | 0 | 54 | 1207 |  | 1705 | 3 | 1162 | 202 | 111 |  | 5 | 75 | 218 | 86 | 60 | 0 | 1 | 4890 |
| 14 | 1 | 72 | 1575 |  | 2900 | 5 | 681 | 79 | 43 |  | 3 | 53 | 0 | 7 | 6 | 0 | 0 | 5425 |
| 15 | 2 | 326 | 7505 |  | 14020 | 26 | 6927 | 308 | 463 |  | 17 | 262 | 79 | 126 | 76 | 0 | 2 | 30140 |
| sum | 27 | 4254 | 36405 |  | 150425 | 234 | 59697 | 2458 | 6197 | 19969 | 337 | 1659 | 24577 | 7409 | 6149 | 0 | 34 | 319831 |

Table 7.2.4.2. Western horse mackerel. Catch-at-age (thousands).

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | ${ }^{15+}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | ${ }^{3713}$ | 21072 | ${ }^{134743}$ | 11515 | 13197 | ${ }^{11741}$ | 8848 | 1651 | 414 | 1651 | 6582 | 18483 | 28679 | 19432 | 8210 |
| 1983 | 0 | 7903 | 2269 | 32900 | 53508 | 15345 | 44539 | 52673 | 17923 | 3291 | 5505 | 3386 | 17017 | 23902 | 38552 | 46482 |
| 1984 | 0 | 0 | 241360 | 4439 | 36294 | 149798 | 22350 | 3824 | 34220 | 14756 | 4101 | 0 | 639 | 1757 | 5080 | 50895 |
| 1985 | 0 | 1633 | 4901 | 60299 | 4463 | 41822 | 100376 | 12644 | 16172 | 6200 | 9224 | 339 | 850 | 3723 | 1250 | 34814 |
| 1986 | 0 | 0 | 0 | 1548 | 676208 | 8727 | 65147 | 10974 | 25712 | 21179 | 15271 | 3116 | 1031 | 855 | 292 | 51531 |
| 1987 | 0 | 99 | 493 | 0 | 2950 | 891600 | 2061 | 41564 | 90814 | 11740 | 9549 | 19663 | 8917 | 1398 | 200 | 32899 |
| 198 | 876 | 27369 | 6112 | 2099 | 4402 | 18968 | 941725 | 12115 | 39913 | 67869 | 9739 | 16326 | 17304 | 5179 | 4892 | 32396 |
| 1989 | 0 | 0 | 0 | 20766 | 18282 | 5308 | 14500 | 1276731 | 12046 | 59357 | 83125 | 13905 | 24196 | 13731 | 8987 | 18132 |
| 199 | 0 | 20406 | 45036 | 138929 | 61442 | 33298 | 10449 | 20607 | 1384850 | 37011 | 70512 | 101945 | 14987 | 34687 | 18077 | 56598 |
| 1991 | 24021 | 20176 | 56066 | 17977 | 159643 | 97147 | 49515 | 21713 | 17148 | 102819 | 20309 | 12161 | ${ }^{43665}$ | 8141 | 7053 | 25553 |
| 192 | 22964 | 14888 | 36332 | 80550 | 56280 | 255874 | 126816 | 48711 | 18992 | 23447 | 1099780 | 13409 | 23002 | 65250 | 11967 | 33246 |
| 1993 | 131108 | 46 | 109807 | 16738 | 62342 | 105760 | 325674 | 141148 | 68418 | 55289 | 30689 | 1075607 | 11373 | 24018 | 68137 | 32140 |
| 1994 | 60759 | 3686 | 911713 | 115729 | 53056 | 44520 | 38769 | 221863 | 106390 | 40988 | 43083 | 22380 | 918512 | 10143 | 14599 | 36635 |
| 1995 | 233030 | 2702 | 646753 | 526053 | 269658 | 74592 | 114649 | 36076 | 22667 | 113304 | 96624 | 59874 | 63187 | 951901 | 39278 | 148243 |
| 1996 | 19774 | 10729 | 659641 | 864188 | 18973 | 87562 | 52050 | 55914 | 58835 | 57361 | 56962 | 91690 | 67114 | 56012 | 34986 | 165611 |
| 1997 | 110451 | 4860 | 471611 | 732959 | 408648 | 256563 | 14168 | 143166 | 143770 | 123043 | 133165 | 96059 | 176730 | 98196 | 51674 | 283111 |
| 1998 | 91505 | 744 | 18443 | 488661 | 359590 | 217571 | 153136 | 119309 | 77494 | 67073 | 50108 | 58791 | 30536 | 65838 | 57584 | 141361 |
| 1999 | 97561 | 14822 | 83714 | 176919 | 265820 | 254516 | 212217 | 187195 | 147271 | 77622 | 35582 | 22909 | 34440 | 2974 | 41831 | 122176 |
| 2000 | 565 | 66210 | 130897 | 64801 | 119297 | 232346 | 202175 | 165745 | 109218 | 54365 | 14594 | 17509 | 18642 | 18585 | 10031 | 73174 |
| 2001 | 60561 | 93125 | 204360 | 166641 | 113659 | 120410 | 141419 | 259974 | 21802 | 110319 | 38576 | 22749 | 17102 | 14092 | 18857 | ${ }_{6} 6868$ |
| 2002 | 14044 | 505717 | 122603 | 158114 | ${ }^{123258}$ | 66640 | 68890 | 95052 | 132743 | 87285 | 46167 | 29692 | 25333 | 11305 | 12753 | 72682 |
| 2003 | 1913 | 323194 | 509889 | 141442 | 148989 | 89122 | 59047 | 48582 | 52305 | 10288 | 57089 | 31778 | 27158 | 8832 | 7683 | 40641 |
| 2004 | 22237 | 159011 | 116055 | 486195 | 81099 | 98855 | 6941 | 48969 | 32589 | 51953 | 54542 | 33298 | 12581 | 13407 | 4305 | 21278 |
| 2005 | 1305 | 74538 | 171420 | 310767 | 540649 | 69957 | 74746 | 61889 | 4443 | 22726 | 2719 | 42746 | 23677 | 6849 | 7491 | 18626 |
| 2006 | 1905 | 53322 | 58091 | 7505 | 91274 | 48229 | 57377 | 37222 | 41970 | 16865 | 11828 | 17073 | 32025 | 12877 | 7464 | 24645 |
| 2007 | 5121 | 32399 | 38598 | 40530 | 61938 | 112724 | 347284 | 48160 | 29112 | 21504 | 8728 | 7015 | 8462 | 14021 | 7618 | 18335 |
| 2008 | 30155 | 78121 | 2446 | 5325 | 57125 | 84358 | 54701 | 297879 | 49889 | 36992 | 25172 | 14466 | 12787 | 9269 | 13194 | 24124 |
| 2009 | 47421 | 86053 | 31431 | 56816 | 40104 | 36174 | 62700 | 57683 | 273217 | 68318 | 42063 | 30583 | 21230 | 8266 | 6811 | 39752 |
| 2010 | 4331 | 68198 | 122386 | 69381 | 29371 | 30496 | 51312 | 110033 | 73973 | 285881 | 70041 | 34486 | 24421 | 14887 | 14942 | 44201 |
| 2011 | 1136 | 17035 | 61864 | 106032 | 51259 | 35380 | 38626 | 59428 | 59031 | 61017 | 239472 | 88764 | 29187 | 17731 | 9783 | 35379 |
| 2012 | 5383 | 48396 | 42933 | 6404 | 171353 | 56060 | 37949 | 28163 | 25641 | 45516 | 41305 | 162155 | 50561 | 24067 | 11649 | 30636 |
| 2013 | 94165 | 138663 | 34651 | 34171 | 76847 | 248958 | 67370 | 25070 | 18447 | 20746 | 31217 | 20836 | 10642 | 21316 | 16279 | 24536 |
| 2014 | 19215 | 26080 | 83034 | 34591 | 28200 | 62102 | 152650 | 56679 | 21786 | 16441 | 23876 | 23654 | 24509 | 57284 | 25197 | 23878 |
| 2015 | 85629 | 108174 | 25416 | 51631 | 31604 | 24613 | 46201 | 118679 | 27331 | 12698 | 10883 | 12584 | 11794 | 7272 | 48586 | 15935 |
| 2016 | 133936 | 168323 | 97368 | 18662 | 31033 | 18762 | 14519 | 22754 | 80818 | 19004 | 10531 | 10298 | 14703 | 16212 | 18451 | 62769 |
| 2017 | 2453 | 3610 | 5217 | 112644 | 20936 | 37499 | 8531 | 7307 | 17117 | 48728 | 8823 | 3471 | 3040 | 4890 | 5425 | 30140 |

Table 7.2.5.1. Western horse mackerel stock. Mean weight (kg) in catch-at-age by quarter and area in 2017

| Q1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.5.b.1.b | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | mean |
| 0 |  |  |  |  |  |  |  |  | 0.015 |  |  |  |  | 0.015 |
| 1 |  |  |  |  |  |  |  |  | 0.041 |  |  |  |  | 0.041 |
| 2 | 0.238 | 0.238 | 0.280 | 0.289 |  |  | 0.052 | 0.046 | 0.092 | 0.051 | 0.052 | 0.052 | 0.052 | 0.131 |
| 3 | 0.101 | 0.101 | 0.087 | 0.084 | 0.078 |  | 0.075 | 0.064 | 0.111 | 0.070 | 0.071 | 0.075 | 0.075 | 0.083 |
| 4 | 0.144 | 0.144 | 0.133 | 0.131 | 0.160 |  | 0.126 | 0.042 | 0.115 | 0.102 | 0.103 | 0.126 | 0.126 | 0.121 |
| 5 | 0.200 | 0.200 | 0.196 | 0.194 | 0.214 | 0.230 | 0.210 |  | 0.181 | 0.156 | 0.156 | 0.210 | 0.210 | 0.197 |
| 6 | 0.220 | 0.220 | 0.215 | 0.214 | 0.238 | 0.259 | 0.232 |  | 0.338 | 0.181 | 0.181 | 0.232 | 0.232 | 0.230 |
| 7 | 0.236 | 0.236 | 0.228 | 0.228 | 0.249 |  | 0.252 |  |  | 0.228 | 0.228 | 0.252 | 0.252 | 0.239 |
| 8 | 0.268 | 0.268 | 0.264 | 0.264 | 0.246 | 0.253 | 0.249 |  | 0.407 | 0.237 | 0.237 | 0.248 | 0.248 | 0.266 |
| 9 | 0.268 | 0.268 | 0.264 | 0.264 | 0.262 | 0.298 | 0.266 |  | 0.387 | 0.248 | 0.248 | 0.265 | 0.265 | 0.275 |
| 10 | 0.297 | 0.297 | 0.287 | 0.285 | 0.277 | 0.280 | 0.281 |  | 0.417 |  |  | 0.276 | 0.276 | 0.297 |
| 11 | 0.324 | 0.324 | 0.319 | 0.323 | 0.324 | 0.296 | 0.338 |  | 0.439 | 0.291 | 0.291 | 0.324 | 0.324 | 0.326 |
| 12 | 0.325 | 0.325 | 0.314 | 0.313 | 0.368 | 0.289 | 0.364 |  | 0.433 | 0.331 | 0.331 | 0.351 | 0.351 | 0.341 |
| 13 | 0.316 | 0.316 | 0.314 | 0.312 | 0.332 | 0.353 | 0.336 |  | 0.408 | 0.306 | 0.306 | 0.334 | 0.334 | 0.331 |
| 14 | 0.332 | 0.332 | 0.326 | 0.324 | 0.345 | 0.321 | 0.343 |  | 0.441 |  |  | 0.331 | 0.331 | 0.342 |
| 15+ | 0.333 | 0.334 | 0.327 | 0.326 | 0.324 | 0.361 | 0.324 |  | 0.427 | 0.374 | 0.374 | 0.320 | 0.320 | 0.346 |


| Q2 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.5.b.1.b | 27.6.a | 27.7.b | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k. 2 | mean |
| 0 |  |  |  |  | 0.015 | 0.015 |  |  |  |  |  | 0.015 |
| 1 |  |  |  |  | 0.041 | 0.041 |  |  |  |  |  | 0.041 |
| 2 | 0.238 | 0.238 |  | 0.052 | 0.084 | 0.092 | 0.051 | 0.051 | 0.051 | 0.051 | 0.051 | 0.096 |
| 3 | 0.101 | 0.101 |  | 0.075 | 0.094 | 0.111 | 0.070 | 0.070 | 0.070 | 0.070 | 0.070 | 0.083 |
| 4 | 0.144 | 0.144 |  | 0.126 | 0.055 | 0.115 | 0.102 | 0.102 | 0.102 | 0.102 | 0.102 | 0.109 |
| 5 | 0.200 | 0.200 |  | 0.210 | 0.163 | 0.181 | 0.156 | 0.156 | 0.156 | 0.156 | 0.156 | 0.174 |
| 6 | 0.220 | 0.220 |  | 0.232 |  | 0.338 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 | 0.213 |
| 7 | 0.236 | 0.260 | 0.269 | 0.252 |  |  | 0.228 | 0.228 | 0.228 | 0.228 | 0.228 | 0.239 |
| 8 | 0.268 | 0.314 | 0.332 | 0.249 |  | 0.407 | 0.237 | 0.237 | 0.237 | 0.237 | 0.237 | 0.276 |
| 9 | 0.268 | 0.307 | 0.323 | 0.266 |  | 0.387 | 0.248 | 0.248 | 0.248 | 0.248 | 0.248 | 0.279 |
| 10 | 0.297 | 0.351 | 0.372 | 0.281 |  | 0.417 |  | 0.270 | 0.270 | 0.270 | 0.270 | 0.311 |
| 11 | 0.324 | 0.363 | 0.379 | 0.338 |  | 0.439 | 0.291 | 0.291 | 0.291 | 0.291 | 0.291 | 0.330 |
| 12 | 0.325 | 0.367 | 0.384 | 0.364 |  | 0.433 | 0.331 | 0.331 | 0.331 | 0.331 | 0.331 | 0.353 |
| 13 | 0.316 | 0.356 | 0.372 | 0.336 |  | 0.408 | 0.306 | 0.306 | 0.306 | 0.306 | 0.306 | 0.332 |
| 14 | 0.332 | 0.346 | 0.352 | 0.343 |  | 0.441 |  | 0.350 | 0.350 | 0.350 | 0.350 | 0.357 |
| 15+ | 0.334 | 0.383 | 0.402 | 0.324 |  | 0.427 | 0.374 | 0.373 | 0.373 | 0.373 | 0.373 | 0.374 |

Table 7.2.5.1 cont. Western horse mackerel stock. Mean weight (kg) in catch-at-age by quarter and area in 2017

| Q3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.3.a | 27.4.a | 27.5.b.1.b |  | 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h |  |  | 27.7.j | 27.7.k. 2 | mean |
| 0 |  |  |  |  |  |  |  |  | 0.015 | 0.015 | 0.015 |  |  |  |  |  | 0.015 |
| 1 | 0.060 |  |  |  |  |  |  |  | 0.041 | 0.041 | 0.041 |  |  |  |  |  | 0.041 |
| 2 | 0.086 |  |  |  | 0.176 |  | 0.052 | 0.052 | 0.084 | 0.092 | 0.092 |  |  |  |  |  | 0.112 |
| 3 | 0.153 | 0.193 |  |  | 0.114 | 0.111 | 0.075 | 0.075 | 0.094 | 0.111 | 0.111 |  | 0.111 |  | 0.111 | 0.111 | 0.102 |
| 4 | 0.182 | 0.234 |  |  | 0.176 | 0.185 | 0.126 | 0.126 | 0.055 | 0.115 | 0.115 |  | 0.185 |  | 0.185 | 0.185 | 0.145 |
| 5 | 0.203 | 0.234 |  |  | 0.213 | 0.245 | 0.210 | 0.210 | 0.163 | 0.181 | 0.169 |  | 0.245 |  | 0.245 | 0.245 | 0.211 |
| 6 | 0.218 | 0.246 |  |  | 0.240 | 0.235 | 0.232 | 0.232 |  | 0.338 |  |  | 0.235 |  | 0.235 | 0.235 | 0.245 |
| 7 | 0.227 | 0.279 |  |  | 0.256 |  | 0.252 | 0.252 |  |  |  |  |  |  |  |  | 0.249 |
| 8 | 0.244 | 0.300 |  |  | 0.282 | 0.317 | 0.249 | 0.249 |  | 0.407 | 0.347 |  | 0.317 |  | 0.317 | 0.317 | 0.307 |
| 9 | 0.249 | 0.277 |  |  | 0.282 | 0.301 | 0.266 | 0.266 |  | 0.387 | 0.342 |  | 0.301 |  | 0.301 | 0.301 | 0.301 |
| 10 | 0.266 | 0.300 |  |  | 0.317 | 0.343 | 0.281 | 0.281 |  | 0.417 | 0.351 |  | 0.343 |  | 0.343 | 0.343 | 0.331 |
| 11 | 0.295 | 0.333 |  |  | 0.326 |  | 0.338 | 0.338 |  | 0.439 | 0.348 |  |  |  |  |  | 0.352 |
| 12 | 0.310 | 0.310 |  |  | 0.352 | 0.400 | 0.364 | 0.364 |  | 0.433 | 0.306 |  | 0.400 |  | 0.400 | 0.400 | 0.374 |
| 13 | 0.309 | 0.309 |  |  | 0.334 | 0.427 | 0.336 | 0.336 |  | 0.408 | 0.359 |  | 0.427 |  | 0.427 | 0.427 | 0.380 |
| 14 | 0.273 | 0.332 |  |  | 0.347 |  | 0.343 | 0.343 |  | 0.441 |  |  |  |  |  |  | 0.361 |
| 15+ | 0.296 | 0.333 |  |  | 0.355 | 0.396 | 0.324 | 0.324 |  | 0.427 | 0.376 |  | 0.396 |  | 0.396 | 0.396 | 0.372 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Q4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Age | 27.2.a | 27.3.a | 27.4.a | 27.5.b.1.b | b 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h |  | 27.7.j | 27.7.k | 27.7.k. 2 | mean |
| 0 |  |  |  |  |  |  |  |  | 0.015 | 0.015 |  |  |  |  |  |  | 0.015 |
| 1 |  | 0.060 |  |  |  |  |  |  | 0.042 | 0.041 |  |  |  |  |  |  | 0.048 |
| 2 | 0.238 | 0.086 | 0.151 | 0.238 | 0.176 |  | 0.052 | 0.052 | 0.092 | 0.092 |  | 0.083 |  |  | 0.052 | 0.052 | 0.114 |
| 3 | 0.101 | 0.153 | 0.120 | 0.101 | 0.119 | 0.147 | 0.075 | 0.075 | 0.112 | 0.111 |  | 0.099 |  |  | 0.075 | 0.075 | 0.105 |
| 4 | 0.144 | 0.182 | 0.188 | 0.144 | 0.177 | 0.193 | 0.126 | 0.126 | 0.115 | 0.115 |  |  |  |  | 0.126 | 0.126 | 0.147 |
| 5 | 0.200 | 0.203 | 0.218 | 0.200 | 0.209 | 0.221 | 0.210 | 0.210 | 0.163 | 0.181 | 0.305 |  |  | 0.293 | 0.210 | 0.210 | 0.217 |
| 6 | 0.220 | 0.218 | 0.248 | 0.220 | 0.235 | 0.234 | 0.232 | 0.232 |  | 0.338 | 0.338 |  |  |  | 0.232 | 0.232 | 0.248 |
| 7 | 0.236 | 0.227 | 0.265 | 0.236 | 0.256 | 0.254 | 0.252 | 0.252 |  |  |  |  |  |  | 0.252 | 0.252 | 0.248 |
| 8 | 0.268 | 0.244 | 0.287 | 0.268 | 0.278 | 0.271 | 0.249 | 0.249 |  | 0.407 | 0.407 |  |  | 0.347 | 0.248 | 0.248 | 0.290 |
| 9 | 0.268 | 0.249 | 0.288 | 0.268 | 0.275 | 0.287 | 0.266 | 0.266 |  | 0.387 | 0.387 |  |  | 0.342 | 0.265 | 0.265 | 0.293 |
| 10 | 0.297 | 0.266 | 0.325 | 0.297 | 0.310 | 0.311 | 0.281 | 0.281 |  | 0.417 | 0.417 |  |  | 0.351 | 0.276 | 0.276 | 0.316 |
| 11 | 0.324 | 0.295 | 0.327 | 0.324 | 0.322 | 0.354 | 0.338 | 0.338 |  | 0.439 | 0.439 |  |  | 0.348 | 0.324 | 0.324 | 0.346 |
| 12 | 0.325 | 0.310 | 0.362 | 0.325 | 0.351 | 0.312 | 0.364 | 0.364 |  | 0.433 | 0.433 |  |  | 0.306 | 0.351 | 0.351 | 0.353 |
| 13 | 0.316 | 0.309 | 0.341 | 0.316 | 0.334 | 0.354 | 0.336 | 0.336 |  | 0.408 | 0.408 |  |  | 0.359 | 0.334 | 0.334 | 0.345 |
| 14 | 0.332 | 0.273 | 0.354 | 0.332 | 0.335 | 0.297 | 0.343 | 0.343 |  | 0.441 | 0.441 |  |  |  | 0.331 | 0.331 | 0.346 |
| 15+ | 0.333 | 0.296 | 0.362 | 0.334 | 0.346 | 0.298 | 0.324 | 0.324 |  | 0.427 | 0.427 |  |  | 0.376 | 0.320 | 0.320 | 0.345 |


| Q1-4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.3.a | 27.4.a | 27.5.b.1.b | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k | 27.7.k. 2 | mean |
| 0 |  |  |  |  |  |  |  |  |  | 0.015 | 0.015 | 0.015 |  |  |  |  |  | 0.015 |
| 1 |  | 0.060 |  |  |  |  |  |  |  | 0.042 | 0.041 | 0.041 |  |  |  |  |  | 0.046 |
| 2 | 0.238 | 0.086 | 0.151 | 0.238 | 0.218 | 0.289 |  | 0.052 | 0.052 | 0.079 | 0.092 | 0.069 | 0.061 | 0.051 | 0.052 | 0.052 | 0.052 | 0.114 |
| 3 | 0.101 | 0.153 | 0.162 | 0.101 | 0.104 | 0.084 | 0.109 | 0.075 | 0.075 | 0.093 | 0.111 | 0.088 | 0.086 | 0.081 | 0.073 | 0.075 | 0.086 | 0.097 |
| 4 | 0.144 | 0.182 | 0.214 | 0.144 | 0.158 | 0.131 | 0.175 | 0.126 | 0.126 | 0.069 | 0.115 | 0.108 | 0.125 | 0.128 | 0.114 | 0.126 | 0.138 | 0.137 |
| 5 | 0.200 | 0.203 | 0.227 | 0.200 | 0.205 | 0.194 | 0.223 | 0.218 | 0.210 | 0.163 | 0.181 | 0.199 | 0.180 | 0.219 | 0.183 | 0.210 | 0.204 | 0.201 |
| 6 | 0.220 | 0.218 | 0.247 | 0.220 | 0.228 | 0.214 | 0.236 | 0.243 | 0.232 |  | 0.338 | 0.241 | 0.196 | 0.210 | 0.206 | 0.232 | 0.216 | 0.231 |
| 7 | 0.236 | 0.227 | 0.273 | 0.236 | 0.249 | 0.228 | 0.256 | 0.252 | 0.252 |  |  | 0.228 | 0.228 | 0.238 | 0.240 | 0.252 | 0.240 | 0.242 |
| 8 | 0.268 | 0.244 | 0.295 | 0.268 | 0.283 | 0.264 | 0.284 | 0.251 | 0.249 |  | 0.407 | 0.316 | 0.259 | 0.281 | 0.243 | 0.248 | 0.267 | 0.277 |
| 9 | 0.268 | 0.249 | 0.281 | 0.268 | 0.281 | 0.264 | 0.289 | 0.279 | 0.266 |  | 0.387 | 0.314 | 0.262 | 0.285 | 0.257 | 0.265 | 0.271 | 0.280 |
| 10 | 0.297 | 0.266 | 0.310 | 0.297 | 0.314 | 0.285 | 0.318 | 0.281 | 0.281 |  | 0.417 | 0.381 | 0.302 | 0.304 | 0.273 | 0.276 | 0.296 | 0.306 |
| 11 | 0.324 | 0.295 | 0.330 | 0.324 | 0.331 | 0.323 | 0.349 | 0.321 | 0.338 |  | 0.439 | 0.348 | 0.291 | 0.318 | 0.308 | 0.324 | 0.308 | 0.330 |
| 12 | 0.325 | 0.310 | 0.332 | 0.325 | 0.344 | 0.313 | 0.363 | 0.334 | 0.364 |  | 0.433 | 0.350 | 0.350 | 0.341 | 0.341 | 0.351 | 0.360 | 0.346 |
| 13 | 0.316 | 0.309 | 0.323 | 0.316 | 0.334 | 0.312 | 0.362 | 0.343 | 0.336 |  | 0.408 | 0.350 | 0.339 | 0.346 | 0.320 | 0.334 | 0.356 | 0.338 |
| 14 | 0.332 | 0.273 | 0.341 | 0.332 | 0.337 | 0.324 | 0.333 | 0.334 | 0.343 |  | 0.441 | 0.441 | 0.350 | 0.342 | 0.340 | 0.331 | 0.340 | 0.346 |
| 15+ | 0.333 | 0.296 | 0.345 | 0.334 | 0.351 | 0.326 | 0.348 | 0.339 | 0.324 |  | 0.427 | 0.388 | 0.380 | 0.365 | 0.347 | 0.320 | 0.363 | 0.349 |

Table 7.2.5.2. Western horse mackerel stock. Mean length (cm) in catch-at-age by quarter and area in 2017

| Q1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.5.b.1.b | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | mean |
| 0 |  |  |  |  |  |  |  | 12.8 |  |  |  |  | 12.8 |
| 1 |  |  |  |  |  |  |  | 17.7 |  |  |  |  | 17.7 |
| 2 | 31.4 | 34.5 | 34.5 |  |  | 19.5 | 18.5 | 22.6 | 19.4 | 19.5 | 19.5 | 19.5 | 22.7 |
| 3 | 23.8 | 22.6 | 22.6 | 22.1 |  | 21.7 | 20.8 | 23.9 | 21.3 | 21.3 | 21.7 | 21.7 | 21.7 |
| 4 | 26.7 | 26.0 | 26.1 | 27.9 |  | 25.7 | 18.5 | 24.2 | 24.2 | 24.3 | 25.7 | 25.7 | 24.2 |
| 5 | 29.7 | 29.5 | 29.5 | 30.0 | 30.3 | 29.8 |  | 26.9 | 27.8 | 27.8 | 29.8 | 29.8 | 29.2 |
| 6 | 30.6 | 30.5 | 30.5 | 31.3 | 31.5 | 31.1 |  | 34.1 | 29.1 | 29.1 | 31.0 | 31.0 | 30.5 |
| 7 | 31.3 | 31.0 | 31.1 | 31.8 |  | 32.0 |  |  | 31.4 | 31.4 | 32.0 | 32.0 | 31.5 |
| 8 | 32.6 | 32.5 | 32.5 | 31.8 | 32.1 | 31.9 |  | 36.2 | 31.9 | 31.9 | 31.9 | 31.9 | 32.1 |
| 9 | 32.7 | 32.5 | 32.5 | 32.3 | 33.6 | 32.4 |  | 35.7 | 32.7 | 32.7 | 32.4 | 32.4 | 32.7 |
| 10 | 33.7 | 33.4 | 33.4 | 32.5 | 32.4 | 32.6 |  | 36.5 |  |  | 32.5 | 32.5 | 33.0 |
| 11 | 34.7 | 34.5 | 34.7 | 34.6 | 33.5 | 34.9 |  | 37.2 | 33.5 | 33.5 | 34.5 | 34.5 | 34.2 |
| 12 | 34.6 | 34.3 | 34.3 | 36.1 | 33.0 | 36.1 |  | 37.0 | 38.5 | 38.5 | 35.7 | 35.7 | 36.1 |
| 13 | 34.5 | 34.4 | 34.4 | 35.0 | 34.9 | 35.0 |  | 36.3 | 34.5 | 34.5 | 34.9 | 34.9 | 34.7 |
| 14 | 34.9 | 34.7 | 34.7 | 35.4 | 34.5 | 35.2 |  | 37.2 |  |  | 34.9 | 34.9 | 35.0 |
| 15+ | 35.0 | 34.9 | 34.9 | 34.5 | 35.7 | 34.5 |  | 36.8 | 37.5 | 37.5 | 34.4 | 34.4 | 35.6 |


| Q2 |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.5.b.1.b | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | mean |
| 0 |  |  |  | 12.8 | 12.8 |  |  |  |  |  | 12.8 |
| 1 |  |  |  | 17.7 | 17.7 |  |  |  |  |  | 17.7 |
| 2 |  |  | 19.5 | 21.9 | 22.6 | 19.4 | 19.4 | 19.4 | 19.4 | 19.4 | 20.1 |
| 3 |  |  | 21.7 | 22.7 | 23.9 | 21.3 | 21.3 | 21.3 | 21.3 | 21.3 | 21.8 |
| 4 |  |  | 25.7 | 19.5 | 24.2 | 24.2 | 24.2 | 24.2 | 24.2 | 24.2 | 23.8 |
| 5 |  |  | 29.8 | 26.0 | 26.9 | 27.8 | 27.8 | 27.8 | 27.8 | 27.8 | 27.7 |
| 6 |  |  | 31.1 |  | 34.1 | 29.1 | 29.1 | 29.1 | 29.1 | 29.1 | 30.1 |
| 7 | 32.5 | 32.5 | 32.0 |  |  | 31.4 | 31.4 | 31.4 | 31.4 | 31.4 | 31.7 |
| 8 | 34.5 | 34.5 | 31.9 |  | 36.2 | 31.9 | 31.9 | 31.9 | 31.9 | 31.9 | 33.0 |
| 9 | 34.2 | 34.2 | 32.4 |  | 35.7 | 32.7 | 32.7 | 32.7 | 32.7 | 32.7 | 33.3 |
| 10 | 35.5 | 35.5 | 32.6 |  | 36.5 |  | 32.6 | 32.6 | 32.6 | 32.6 | 33.8 |
| 11 | 35.5 | 35.5 | 34.9 |  | 37.2 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 34.5 |
| 12 | 35.5 | 35.5 | 36.1 |  | 37.0 | 38.5 | 38.4 | 38.4 | 38.4 | 38.4 | 37.3 |
| 13 | 34.5 | 34.5 | 35.0 |  | 36.3 | 34.5 | 34.5 | 34.5 | 34.5 | 34.5 | 34.8 |
| 14 | 35.2 | 35.2 | 35.2 |  | 37.2 |  | 35.7 | 35.7 | 35.7 | 35.7 | 35.7 |
| 15+ | 36.8 | 36.8 | 34.5 |  | 36.8 | 37.5 | 37.5 | 37.5 | 37.5 | 37.5 | 36.9 |

Table 7.2.5.2 cont. Western horse mackerel stock. Mean length (cm) in catch-at-age by quarter and area in 2017

| Q3 | 27.3.a | 27.4.a | 27.6.a |  |  | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j |  | 27.7.k. 2 | mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 |  |  |  |  |  |  |  |  | 12.8 | 12.8 | 12.8 |  |  |  |  | 12.8 |
| 1 | 19.1 |  |  |  |  |  |  |  | 17.7 | 17.7 | 17.7 |  |  |  |  | 18.1 |
| 2 | 21.3 |  |  | 26.1 |  |  | 19.5 | 19.5 | 21.9 | 22.6 | 22.6 |  |  |  |  | 21.9 |
| 3 | 25.9 | 28.5 |  | 25.1 |  | 23.0 | 21.7 | 21.7 | 22.7 | 23.9 | 23.9 | 23.0 |  | 23.0 | 23.0 | 23.8 |
| 4 | 27.2 | 30.4 |  | 28.7 |  | 27.3 | 25.7 | 25.7 | 19.5 | 24.2 | 24.2 | 27.3 |  | 27.3 | 27.3 | 26.2 |
| 5 | 28.6 | 30.7 |  | 30.3 |  | 29.8 | 29.8 | 29.8 | 26.0 | 26.9 | 26.3 | 29.8 |  | 29.8 | 29.8 | 29.0 |
| 6 | 29.5 | 31.5 |  | 31.5 |  | 29.5 | 31.1 | 31.1 |  | 34.1 |  | 29.5 |  | 29.5 | 29.5 | 30.7 |
| 7 | 30.2 | 33.0 |  | 32.3 |  |  | 32.0 | 32.0 |  |  |  |  |  |  |  | 31.9 |
| 8 | 30.8 | 33.6 |  | 33.1 |  | 33.7 | 31.9 | 31.9 |  | 36.2 | 34.1 | 33.7 |  | 33.7 | 33.7 | 33.3 |
| 9 | 31.0 | 32.9 |  | 33.2 |  | 33.3 | 32.4 | 32.4 |  | 35.7 | 34.4 | 33.3 |  | 33.3 | 33.3 | 33.2 |
| 10 | 31.8 | 34.0 |  | 34.5 |  | 34.0 | 32.6 | 32.6 |  | 36.5 | 34.5 | 34.0 |  | 34.0 | 34.0 | 33.9 |
| 11 | 33.3 | 35.5 |  | 34.6 |  |  | 34.9 | 34.9 |  | 37.2 | 34.5 |  |  |  |  | 35.0 |
| 12 | 33.4 | 33.4 |  | 35.7 |  | 35.5 | 36.1 | 36.1 |  | 37.0 | 33.5 | 35.5 |  | 35.5 | 35.5 | 35.2 |
| 13 | 34.0 | 34.0 |  | 35.0 |  | 37.5 | 35.0 | 35.0 |  | 36.3 | 35.0 | 37.5 |  | 37.5 | 37.5 | 35.8 |
| 14 | 32.8 | 35.1 |  | 35.4 |  |  | 35.2 | 35.2 |  | 37.2 |  |  |  |  |  | 35.1 |
| 15+ | 33.5 | 35.4 |  | 35.7 |  | 36.3 | 34.5 | 34.5 |  | 36.8 | 35.5 | 36.3 |  | 36.3 | 36.3 | 35.5 |
| Q4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Age | 27.2.a | 27.3.a | 27.4.a |  | 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.k | 27.7.k. 2 | mean |
| 0 |  |  |  |  |  |  |  |  | 12.8 | 12.8 |  |  |  |  |  | 12.8 |
| 1 |  | 19.1 |  |  |  |  |  |  | 17.9 | 17.7 |  |  |  |  |  | 18.2 |
| 2 | 31.4 | 21.3 | 26.1 |  | 26.1 |  | 19.5 | 19.5 | 22.6 | 22.6 |  | 22.0 |  | 19.5 | 19.5 | 22.7 |
| 3 | 23.8 | 25.9 | 25.1 |  | 25.2 | 26.7 | 21.7 | 21.7 | 23.9 | 23.9 |  | 23.7 |  | 21.7 | 21.7 | 23.7 |
| 4 | 26.7 | 27.2 | 28.7 |  | 28.6 | 29.2 | 25.7 | 25.7 | 24.2 | 24.2 |  |  |  | 25.7 | 25.7 | 26.5 |
| 5 | 29.7 | 28.6 | 30.3 |  | 30.0 | 30.4 | 29.8 | 29.8 | 26.0 | 26.9 | 33.0 |  | 32.2 | 29.8 | 29.8 | 29.7 |
| 6 | 30.6 | 29.5 | 31.5 |  | 31.1 | 31.0 | 31.1 | 31.1 |  | 34.1 | 34.1 |  |  | 31.0 | 31.0 | 31.5 |
| 7 | 31.3 | 30.2 | 32.3 |  | 32.2 | 31.8 | 32.0 | 32.0 |  |  |  |  |  | 32.0 | 32.0 | 31.7 |
| 8 | 32.6 | 30.8 | 33.1 |  | 32.8 | 32.4 | 31.9 | 31.9 |  | 36.2 | 36.2 |  | 34.1 | 31.9 | 31.9 | 33.0 |
| 9 | 32.7 | 31.0 | 33.2 |  | 32.7 | 33.0 | 32.4 | 32.4 |  | 35.7 | 35.7 |  | 34.4 | 32.4 | 32.4 | 33.2 |
| 10 | 33.7 | 31.8 | 34.4 |  | 34.0 | 33.9 | 32.6 | 32.6 |  | 36.5 | 36.5 |  | 34.5 | 32.5 | 32.5 | 33.8 |
| 11 | 34.7 | 33.3 | 34.5 |  | 34.4 | 35.4 | 34.9 | 34.9 |  | 37.2 | 37.2 |  | 34.5 | 34.5 | 34.5 | 35.0 |
| 12 | 34.6 | 33.4 | 35.7 |  | 35.7 | 33.9 | 36.1 | 36.1 |  | 37.0 | 37.0 |  | 33.5 | 35.7 | 35.7 | 35.4 |
| 13 | 34.5 | 34.0 | 35.0 |  | 35.0 | 35.4 | 35.0 | 35.0 |  | 36.3 | 36.3 |  | 35.0 | 34.9 | 34.9 | 35.1 |
| 14 | 34.9 | 32.8 | 35.4 |  | 34.7 | 33.5 | 35.2 | 35.2 |  | 37.2 | 37.2 |  |  | 34.9 | 34.9 | 35.1 |
| 15+ | 35.0 | 33.5 | 35.7 |  | 35.3 | 33.5 | 34.5 | 34.5 |  | 36.8 | 36.8 |  | 35.5 | 34.4 | 34.4 | 35.0 |

Table 7.2.5.2 cont. Western horse mackerel stock. Mean length (cm) in catch-at-age by quarter and area in 2017

| Q1-4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.3.a | 27.4.a | 27.5.b.1.b | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k | 27.7.k. 2 | mean |
| 0 |  |  |  |  |  |  |  |  |  | 12.8 | 12.8 | 12.8 |  |  |  |  |  | 12.8 |
| 1 |  | 19.1 |  |  |  |  |  |  |  | 17.8 | 17.7 | 17.7 |  |  |  |  |  | 18.1 |
| 2 | 31.4 | 21.3 | 26.1 | 31.4 | 30.2 | 34.5 |  | 19.5 | 19.5 | 21.4 | 22.6 | 20.8 | 20.2 | 19.4 | 19.5 | 19.5 | 19.5 | 23.5 |
| 3 | 23.8 | 25.9 | 27.0 | 23.8 | 23.9 | 22.6 | 23.8 | 21.7 | 21.7 | 22.7 | 23.9 | 22.4 | 22.2 | 21.8 | 21.5 | 21.7 | 22.0 | 23.1 |
| 4 | 26.7 | 27.2 | 29.7 | 26.7 | 27.3 | 26.1 | 28.2 | 25.7 | 25.7 | 20.6 | 24.2 | 24.2 | 25.1 | 25.4 | 24.9 | 25.7 | 25.7 | 25.8 |
| 5 | 29.7 | 28.6 | 30.5 | 29.7 | 29.8 | 29.5 | 30.1 | 30.0 | 29.8 | 26.0 | 26.9 | 28.7 | 28.4 | 29.7 | 28.8 | 29.8 | 29.2 | 29.1 |
| 6 | 30.6 | 29.5 | 31.5 | 30.6 | 30.9 | 30.5 | 30.8 | 31.2 | 31.1 |  | 34.1 | 31.1 | 29.2 | 29.9 | 30.1 | 31.0 | 29.9 | 30.7 |
| 7 | 31.3 | 30.2 | 32.7 | 31.3 | 31.8 | 31.1 | 32.0 | 32.0 | 32.0 |  |  | 31.4 | 31.4 | 31.6 | 31.7 | 32.0 | 31.7 | 31.6 |
| 8 | 32.6 | 30.8 | 33.4 | 32.6 | 33.0 | 32.5 | 32.9 | 32.0 | 31.9 |  | 36.2 | 33.7 | 32.4 | 32.8 | 31.9 | 31.9 | 32.5 | 32.7 |
| 9 | 32.7 | 31.0 | 33.0 | 32.7 | 33.0 | 32.5 | 33.1 | 32.9 | 32.4 |  | 35.7 | 34.0 | 32.9 | 33.2 | 32.6 | 32.4 | 32.8 | 32.9 |
| 10 | 33.7 | 31.8 | 34.2 | 33.7 | 34.1 | 33.4 | 33.8 | 32.6 | 32.6 |  | 36.5 | 35.4 | 33.2 | 33.3 | 32.5 | 32.5 | 33.0 | 33.5 |
| 11 | 34.7 | 33.3 | 35.1 | 34.7 | 34.7 | 34.7 | 35.1 | 34.3 | 34.9 |  | 37.2 | 34.8 | 33.5 | 34.1 | 34.0 | 34.5 | 34.0 | 34.6 |
| 12 | 34.6 | 33.4 | 34.4 | 34.6 | 35.0 | 34.3 | 35.3 | 34.8 | 36.1 |  | 37.0 | 36.5 | 37.6 | 36.0 | 37.0 | 35.7 | 36.5 | 35.6 |
| 13 | 34.5 | 34.0 | 34.4 | 34.5 | 34.6 | 34.4 | 35.4 | 35.0 | 35.0 |  | 36.3 | 35.1 | 35.3 | 35.2 | 34.7 | 34.9 | 35.7 | 34.9 |
| 14 | 34.9 | 32.8 | 35.2 | 34.9 | 34.9 | 34.7 | 34.8 | 34.9 | 35.2 |  | 37.2 | 37.2 | 35.7 | 35.3 | 35.3 | 34.9 | 35.3 | 35.2 |
| ${ }^{15+}$ | 35.0 | 33.5 | 35.5 | 35.1 | 35.4 | 34.9 | 35.1 | 34.9 | 34.5 |  | 36.8 | 36.7 | 37.1 | 36.0 | 35.9 | 34.4 | 36.0 | 35.4 |

Table 7.2.5.3. Western horse mackerel. Catch weights-at-age (kg).

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0.015 | 0.054 | 0.09 | 0.142 | 0.178 | 0.227 | 0.273 | 0.276 | 0.292 | 0.305 | 0.369 | 0.348 | 0.348 | 0.348 | 0.356 | 0.366 |
| 1983 | 0.015 | 0.039 | 0.113 | 0.124 | 0.168 | 0.229 | 0.247 | 0.282 | 0.281 | 0.254 | 0.26 | 0.3 | 0.31 | 0.315 | 0.311 | 0.332 |
| 1984 | 0.015 | 0.034 | 0.073 | 0.089 | 0.13 | 0.176 | 0.216 | 0.245 | 0.278 | 0.262 | 0.259 | 0.255 | 0.344 | 0.232 | 0.306 | 0.308 |
| 1985 | 0.015 | 0.029 | 0.045 | 0.087 | 0.15 | 0.156 | 0.199 | 0.243 | 0.256 | 0.294 | 0.257 | 0.241 | 0.251 | 0.314 | 0.346 | 0.321 |
| 1986 | 0.015 | 0.029 | 0.045 | 0.11 | 0.107 | 0.171 | 0.196 | 0.223 | 0.251 | 0.296 | 0.28 | 0.319 | 0.287 | 0.345 | 0.26 | 0.36 |
| 1987 | 0.015 | 0.068 | 0.067 | 0.11 | 0.155 | 0.143 | 0.174 | 0.198 | 0.249 | 0.264 | 0.321 | 0.336 | 0.244 | 0.328 | 0.245 | 0.373 |
| 1988 | 0.015 | 0.031 | 0.075 | 0.114 | 0.132 | 0.147 | 0.157 | 0.24 | 0.304 | 0.335 | 0.386 | 0.434 | 0.404 | 0.331 | 0.392 | 0.424 |
| 1989 | 0.012 | 0.05 | 0.075 | 0.149 | 0.142 | 0.142 | 0.22 | 0.166 | 0.258 | 0.327 | 0.33 | 0.381 | 0.4 | 0.421 | 0.448 | 0.516 |
| 1990 | 0.015 | 0.032 | 0.031 | 0.09 | 0.124 | 0.126 | 0.129 | 0.202 | 0.183 | 0.227 | 0.32 | 0.328 | 0.355 | 0.399 | 0.388 | 0.379 |
| 1991 | 0.012 | 0.031 | 0.046 | 0.113 | 0.125 | 0.148 | 0.141 | 0.144 | 0.187 | 0.185 | 0.215 | 0.303 | 0.323 | 0.354 | 0.365 | 0.33 |
| 1992 | 0.008 | 0.014 | 0.092 | 0.117 | 0.139 | 0.143 | 0.157 | 0.163 | 0.172 | 0.235 | 0.222 | 0.288 | 0.306 | 0.359 | 0.393 | 0.401 |
| 1993 | 0.01 | 0.033 | 0.083 | 0.12 | 0.126 | 0.142 | 0.154 | 0.163 | 0.183 | 0.199 | 0.177 | 0.238 | 0.308 | 0.327 | 0.376 | 0.421 |
| 1994 | 0.021 | 0.037 | 0.052 | 0.106 | 0.124 | 0.158 | 0.153 | 0.167 | 0.194 | 0.199 | 0.28 | 0.275 | 0.24 | 0.326 | 0.342 | 0.383 |
| 1995 | 0.015 | 0.038 | 0.052 | 0.073 | 0.089 | 0.126 | 0.13 | 0.17 | 0.176 | 0.2 | 0.204 | 0.222 | 0.215 | 0.246 | 0.237 | 0.298 |
| 1996 | 0.015 | 0.059 | 0.078 | 0.09 | 0.125 | 0.141 | 0.155 | 0.166 | 0.177 | 0.191 | 0.206 | 0.224 | 0.233 | 0.229 | 0.28 | 0.332 |
| 1997 | 0.017 | 0.039 | 0.075 | 0.093 | 0.109 | 0.142 | 0.179 | 0.189 | 0.199 | 0.209 | 0.234 | 0.24 | 0.246 | 0.272 | 0.309 | 0.288 |
| 1998 | 0.014 | 0.041 | 0.087 | 0.102 | 0.113 | 0.14 | 0.162 | 0.172 | 0.183 | 0.192 | 0.213 | 0.227 | 0.242 | 0.231 | 0.239 | 0.272 |
| 1999 | 0 | 0.05 | 0.089 | 0.108 | 0.121 | 0.14 | 0.162 | 0.186 | 0.203 | 0.21 | 0.217 | 0.231 | 0.29 | 0.276 | 0.263 | 0.362 |
| 2000 | 0.026 | 0.058 | 0.084 | 0.101 | 0.118 | 0.149 | 0.164 | 0.155 | 0.193 | 0.209 | 0.234 | 0.215 | 0.252 | 0.195 | 0.274 | 0.362 |
| 2001 | 0.018 | 0.045 | 0.071 | 0.104 | 0.113 | 0.129 | 0.139 | 0.151 | 0.169 | 0.195 | 0.223 | 0.227 | 0.296 | 0.277 | 0.271 | 0.304 |
| 2002 | 0.018 | 0.037 | 0.062 | 0.095 | 0.124 | 0.143 | 0.152 | 0.167 | 0.182 | 0.210 | 0.256 | 0.296 | 0.357 | 0.32 | 0.309 | 0.369 |
| 2003 | 0.041 | 0.062 | 0.061 | 0.082 | 0.127 | 0.149 | 0.172 | 0.167 | 0.178 | 0.189 | 0.220 | 0.275 | 0.363 | 0.315 | 0.331 | 0.379 |
| 2004 | 0.033 | 0.037 | 0.091 | 0.097 | 0.136 | 0.152 | 0.164 | 0.173 | 0.190 | 0.179 | 0.192 | 0.208 | 0.355 | 0.257 | 0.366 | 0.33 |
| 2005 | 0.022 | 0.030 | 0.059 | 0.085 | 0.106 | 0.157 | 0.184 | 0.214 | 0.209 | 0.203 | 0.237 | 0.229 | 0.300 | 0.309 | 0.342 | 0.410 |
| 2006 | 0.036 | 0.039 | 0.062 | 0.095 | 0.131 | 0.135 | 0.191 | 0.225 | 0.258 | 0.291 | 0.271 | 0.256 | 0.269 | 0.269 | 0.319 | 0.380 |
| 2007 | 0.013 | 0.050 | 0.076 | 0.101 | 0.119 | 0.125 | 0.159 | 0.178 | 0.220 | 0.231 | 0.239 | 0.248 | 0.247 | 0.256 | 0.286 | 0.371 |
| 2008 | 0.033 | 0.045 | 0.088 | 0.105 | 0.120 | 0.141 | 0.161 | 0.181 | 0.219 | 0.249 | 0.278 | 0.272 | 0.302 | 0.292 | 0.306 | 0.352 |
| 2009 | 0.031 | 0.041 | 0.072 | 0.095 | 0.121 | 0.160 | 0.177 | 0.219 | 0.226 | 0.279 | 0.311 | 0.396 | 0.404 | 0.308 | 0.355 | 0.446 |
| 2010 | 0.044 | 0.040 | 0.087 | 0.110 | 0.143 | 0.164 | 0.180 | 0.193 | 0.217 | 0.237 | 0.248 | 0.279 | 0.297 | 0.322 | 0.308 | 0.347 |
| 2011 | 0.034 | 0.048 | 0.087 | 0.111 | 0.134 | 0.164 | 0.182 | 0.192 | 0.211 | 0.228 | 0.255 | 0.304 | 0.303 | 0.299 | 0.346 | 0.363 |
| 2012 | 0.058 | 0.061 | 0.083 | 0.108 | 0.148 | 0.166 | 0.182 | 0.207 | 0.228 | 0.242 | 0.266 | 0.269 | 0.296 | 0.298 | 0.336 | 0.375 |
| 2013 | 0.039 | 0.056 | 0.089 | 0.112 | 0.141 | 0.163 | 0.191 | 0.226 | 0.235 | 0.254 | 0.250 | 0.287 | 0.294 | 0.316 | 0.368 | 0.388 |
| 2014 | 0.035 | 0.055 | 0.086 | 0.119 | 0.140 | 0.172 | 0.197 | 0.223 | 0.250 | 0.258 | 0.271 | 0.289 | 0.301 | 0.304 | 0.342 | 0.374 |
| 2015 | 0.020 | 0.015 | 0.075 | 0.113 | 0.139 | 0.162 | 0.196 | 0.220 | 0.246 | 0.263 | 0.279 | 0.312 | 0.298 | 0.317 | 0.322 | 0.339 |
| 2016 | 0.016 | 0.024 | 0.076 | 0.110 | 0.153 | 0.177 | 0.193 | 0.228 | 0.241 | 0.262 | 0.286 | 0.294 | 0.311 | 0.323 | 0.327 | 0.34 |
| 2017 | 0.013 | 0.013 | 0.055 | 0.097 | 0.136 | 0.193 | 0.196 | 0.223 | 0.257 | 0.268 | 0.289 | 0.312 | 0.328 | 0.324 | 0.339 | 0.344 |

Table 7.2.6.1. Western horse mackerel. Maturity-at-age.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1983 | 0 | 0 | 0.3 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1984 | 0 | 0 | 0.1 | 0.6 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1985 | 0 | 0 | 0.1 | 0.4 | 0.8 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1993 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1994 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1995 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1996 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1997 | 0 | 0 | 0.1 | 0.4 | 0.6 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1998 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1999 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2000 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2001 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2002 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2003 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2004 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2005 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2006 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2007 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2008 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2009 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2010 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2011 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2012 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2013 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2014 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2015 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2016 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2017 | 0 | 0 | 0.05 | 0.25 | 0.7 | 0.95 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 7.2.8.1. Western horse mackerel. Potential fecundity ( $10^{6}$ eggs) per kg spawning female vs. weight in kg.

|  | 1987 |  | 1992 |  | 1995 |  | 1998 |  | 2000 |  | 2001 |  | 2001 (CONT) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | w | pfec. | w | pfec. | w | pfec. | w | pfec. | w | pfec. | w | pfec. | w | pfec. |
| 1 | 0.168 | 1.524 | 0.105 | 1.317 | 0.13 | 1.307 | 0.172 | 1.318 | 0.258 | 0.841 | 0.086 | 0.688 | 0.165 | 1.382 |
| 2 | 0.179 | 0.916 | 0.109 | 2.056 | 0.157 | 1.246 | 0.104 | 0.867 | 0.268 | 0.747 | 0.08 | 0.812 | 0.166 | 1.579 |
| 3 | 0.192 | 2.083 | 0.11 | 1.869 | 0.168 | 1.699 | 0.112 | 1.312 | 0.304 | 1.188 | 0.081 | 0.535 | 0.167 | 1.479 |
| 4 | 0.233 | 1.644 | 0.112 | 1.772 | 0.179 | 1.135 | 0.206 | 0.382 | 0.311 | 1.411 | 0.095 | 0.88 | 0.113 | 0.527 |
| 5 | 0.213 | 1.066 | 0.115 | 1.188 | 0.189 | 1.529 | 0.207 | 0.78 | 0.337 | 0.613 | 0.11 | 1.164 | 0.14 | 0.876 |
| 6 | 0.217 | 2.392 | 0.119 | 1.317 | 0.168 | 1.1 | 0.109 | 1.133 | 0.339 | 1.571 | 0.113 | 1.106 | 0.122 | 0.589 |
| 7 | 0.277 | 1.617 | 0.12 | 1.413 | 0.209 | 1.497 | 0.132 | 1.02 | 0.341 | 1.522 | 0.095 | 0.823 | 0.12 | 0.68 |
| 8 | 0.279 | 1.018 | 0.123 | 1.293 | 0.215 | 1.524 | 0.2 | 1.088 | 0.355 | 1.056 | 0.11 | 0.883 | 0.121 | 0.578 |
| 9 | 0.274 | 1.62 | 0.123 | 1.991 | 0.218 | 1.616 | 0.152 | 1.417 | 0.357 | 0.604 | 0.108 | 0.823 | 0.139 | 0.723 |
| 10 | 0.3 | 1.513 | 0.131 | 1.617 | 0.226 | 1.883 | 0.149 | 1.004 | 0.367 | 1.15 | 0.097 | 0.741 | 0.144 | 1.213 |
| 11 | 0.32 | 1.647 | 0.135 | 0.793 | 0.22 | 1.324 |  |  | 0.393 | 1.279 | 0.101 | 0.853 | 0.144 | 1.265 |
| 12 | 0.273 | 1.956 | 0.131 | 1.039 | 0.236 | 1.221 |  |  | 0.393 | 0.668 | 0.106 | 1.133 | 0.171 | 0.956 |
| 13 | 0.212 | 2.83 | 0.136 | 1.06 | 0.261 | 1.21 |  |  | 0.413 | 0.694 | 0.107 | 0.935 | 0.121 | 0.607 |
| 14 | 0.268 | 1.687 | 0.138 | 1.489 | 0.245 | 1.445 |  |  | 0.421 | 1.339 | 0.107 | 0.494 | 0.122 | 0.689 |
| 15 | 0.32 | 1.088 | 0.147 | 1.214 | 0.306 | 1.693 |  |  | 0.423 | 0.798 | 0.11 | 0.85 | 0.139 | 0.915 |
| 16 | 0.318 | 1.208 | 0.151 | 1.158 | 0.314 | 1.312 |  |  | 0.445 | 1.03 | 0.111 | 0.67 | 0.153 | 0.943 |
| 17 | 0.343 | 1.933 | 0.16 | 1.349 | 0.46 | 1.575 |  |  | 0.446 | 1.208 | 0.103 | 0.632 | 0.154 | 0.709 |
| 18 | 0.378 | 1.429 | 0.165 | 1.359 | 0.449 | 1.43 |  |  | 0.152 | 0.643 | 0.111 | 0.547 | 0.156 | 0.773 |
| 19 | 0.404 | 1.849 | 0.165 | 0.945 |  |  |  |  | 0.165 | 0.579 | 0.118 | 0.88 | 0.162 | 1.158 |
| 20 | 0.428 | 2.236 | 0.167 | 1 |  |  |  |  | 0.175 | 0.596 | 0.107 | 0.944 | 0.174 | 1.389 |
| 21 | 0.398 | 1.538 | 0.168 | 1.545 |  |  |  |  | 0.179 | 0.997 | 0.104 | 0.724 | 0.175 | 1.426 |
| 22 | 0.431 | 1.223 | 0.18 | 1.299 |  |  |  |  | 0.19 | 0.744 | 0.111 | 0.86 | 0.179 | 1.248 |
| 23 | 0.432 | 1.465 | 0.174 | 1.487 |  |  |  |  | 0.197 | 0.613 | 0.11 | 0.728 | 0.179 | 1.236 |
| 24 | 0.421 | 1.843 | 0.178 | 1.594 |  |  |  |  | 0.203 | 0.702 | 0.111 | 0.544 | 0.18 | 2.353 |
| 25 | 0.481 | 1.757 | 0.185 | 1.475 |  |  |  |  | 0.219 | 0.472 | 0.129 | 0.935 | 0.184 | 2.255 |
| 26 | 0.494 | 1.611 | 0.195 | 1.41 |  |  |  |  | 0.223 | 0.806 | 0.114 | 0.901 | 0.139 | 0.931 |
| 27 | 0.54 | 1.754 | 0.203 | 1.937 |  |  |  |  | 0.227 | 0.606 | 0.114 | 0.557 | 0.161 | 1.037 |
| 28 | 0.564 | 2.255 | 0.205 | 1.534 |  |  |  |  | 0.289 | 1.273 | 0.151 | 1.377 | 0.162 | 0.893 |
| 29 | 0.585 | 1.221 | 0.213 | 1.577 |  |  |  |  | 0.294 | 1.395 | 0.153 | 1.596 | 0.169 | 0.691 |
| 30 |  |  | 0.222 | 0.958 |  |  |  |  | 0.3 | 1.305 | 0.154 | 1.699 | 0.18 | 1.609 |
| 31 |  |  | 0.275 | 2.444 |  |  |  |  |  |  | 0.103 | 0.679 | 0.185 | 1.776 |
| 32 |  |  |  |  |  |  |  |  |  |  | 0.12 | 1.14 | 0.211 | 2.102 |
| 33 |  |  |  |  |  |  |  |  |  |  | 0.12 | 0.631 | 0.224 | 1.466 |
| 34 |  |  |  |  |  |  |  |  |  |  | 0.121 | 0.834 | 0.162 | 0.849 |
| 35 |  |  |  |  |  |  |  |  |  |  | 0.144 | 0.626 | 0.17 | 0.668 |
| 36 |  |  |  |  |  |  |  |  |  |  | 0.116 | 0.668 | 0.187 | 1.453 |
| 37 |  |  |  |  |  |  |  |  |  |  | 0.118 | 1.194 | 0.198 | 1.371 |
| 38 |  |  |  |  |  |  |  |  |  |  | 0.112 | 0.779 | 0.219 | 1.847 |
| 39 |  |  |  |  |  |  |  |  |  |  | 0.126 | 0.782 | 0.22 | 1.578 |
| 40 |  |  |  |  |  |  |  |  |  |  | 0.139 | 1.244 | 0.201 | 0.878 |
| 41 |  |  |  |  |  |  |  |  |  |  | 0.119 | 1.212 | 0.206 | 1.196 |
| 42 |  |  |  |  |  |  |  |  |  |  | 0.109 | 0.755 | 0.223 | 1.115 |
| 43 |  |  |  |  |  |  |  |  |  |  | 0.122 | 0.841 | 0.225 | 1.43 |
| 44 |  |  |  |  |  |  |  |  |  |  | 0.131 | 0.929 | 0.233 | 1.724 |
| 45 |  |  |  |  |  |  |  |  |  |  | 0.135 | 0.862 | 0.241 | 1.131 |
| 46 |  |  |  |  |  |  |  |  |  |  | 0.142 | 1.834 | 0.219 | 0.96 |
| 47 |  |  |  |  |  |  |  |  |  |  | 0.146 | 1.689 | 0.237 | 1.33 |
| 48 |  |  |  |  |  |  |  |  |  |  | 0.148 | 1.357 | 0.241 | 0.918 |
| 49 |  |  |  |  |  |  |  |  |  |  | 0.151 | 1.817 | 0.34 | 0.605 |


|  | 1987 | 1992 | 1995 | 1998 | 2000 | 2001 | 2001 (CONT) |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 50 |  |  |  |  | 0.164 | 1.631 | 0.407 | 1.189 |
| 51 |  |  |  |  | 0.164 | 1.052 |  |  |

Table 7.3.1.1. Western horse mackerel. Final assessment. Numbers-at-age (thousands).

| YEA |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| R | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| 19 | 51838 | 80824 | 26875 | 65764 | 91791 | 13899 | 12833 | 61263 | 24939 | 21611 | 21326 | 37794 | 62431 | 99008 | 46487 | 15363 | 14151 | 1299 | 1190 | 1087 | 12427 |
| 82 | 500 | 7 | 50 | 00 | 6 | 20 | 80 | 9 | 3 | 0 | 1 | 3 | 9 | 7 | 5 | 9 | 1 | 88 | 71 | 67 | 20 |
| 19 | 10858 | 44596 | 69245 | 22918 | 55800 | 77550 | 11706 | 10788 | 51443 | 20929 | 18131 | 17889 | 31700 | 52363 | 83040 | 38989 | 12885 | 1186 | 1090 | 9986 | 11334 |
| 83 | 70 | 000 | 1 | 50 | 30 | 6 | 60 | 50 | 8 | 6 | 0 | 2 | 8 | 9 | 1 | 2 | 6 | 84 | 19 | 3 | 70 |
| 19 | 12663 | 93401 | 38150 | 58877 | 19358 | 46866 | 64871 | 97677 | 89887 | 42829 | 17417 | 15085 | 14883 | 26372 | 43560 | 69078 | 32433 | 1071 | 9872 | 9068 | 10259 |
| 84 | 50 | 3 | 900 | 0 | 50 | 90 | 3 | 7 | 2 | 0 | 9 | 8 | 1 | 2 | 4 | 1 | 3 | 89 | 6 | 6 | 30 |
| 19 | 20268 | 10893 | 79953 | 32479 | 49828 | 16300 | 39319 | 54300 | 81655 | 75091 | 35766 | 14543 | 12594 | 12424 | 22015 | 36363 | 57665 | 2707 | 8947 | 8241 | 93211 |
| 85 | 20 | 20 | 0 | 400 | 0 | 40 | 70 | 8 | 9 | 8 | 8 | 1 | 7 | 8 | 6 | 9 | 1 | 45 | 8 | 4 | 0 |
| 19 | 27218 | 17436 | 93335 | 68195 | 27567 | 42114 | 13735 | 33070 | 45622 | 68566 | 63035 | 30019 | 12205 | 10569 | 10426 | 18475 | 30515 | 4839 | 2272 | 7508 | 85135 |
| 86 | 50 | 70 | 4 | 3 | 300 | 9 | 70 | 70 | 1 | 1 | 8 | 9 | 4 | 8 | 9 | 3 | 9 | 12 | 02 | 7 | 5 |
| 19 | 54852 | 23413 | 14924 | 79439 | 57691 | 23200 | 35311 | 11490 | 27627 | 38086 | 57219 | 52594 | 25045 | 10182 |  |  | 15411 | 2545 | 4036 | 1895 | 77281 |
| 87 | 20 | 30 | 20 | 9 | 4 | 500 | 9 | 10 | 40 | 1 | 6 | 7 | 0 | 2 | 88174 | 86981 | 8 | 57 | 68 | 26 | 0 |
| 19 | 21870 | 47176 | 20009 | 12664 | 66897 | 48264 | 19318 | 29316 | 95232 | 22877 | 31524 | 47350 | 43517 | 20721 |  |  |  | 1275 | 2105 | 3339 | 79611 |
| 88 | 10 | 40 | 90 | 60 | 1 | 6 | 300 | 3 | 3 | 90 | 3 | , | 5 | 2 | 84240 | 72947 | 71959 | 00 | 91 | 46 | 5 |
| 19 | 23580 | 18808 | 40287 | 16954 | 10639 | 55793 | 40044 | 15975 | 24198 | 78530 | 18856 | 25975 | 39010 | 35850 | 17069 |  |  | 5927 | 1050 | 1734 | 93085 |
| 89 | 60 | 20 | 80 | 30 | 80 | 8 | 1 | 600 | 7 | 7 | 10 |  | 6 | 4 | 7 | 69393 | 60090 | 5 | 26 | 70 | 8 |
| 19 | 17797 | 20278 | 16059 | 34126 | 14237 | 88691 | 46262 | 33092 | 13177 | 19940 | 64678 | 15525 | 21385 | 32114 | 29511 | 14051 |  | 4946 | 4879 | 8644 | 90899 |
| 90 | 60 | 90 | 70 | 70 | 80 | 3 | 4 | 9 | 500 | 5 | 4 | 90 | 1 | 0 | 2 | 0 | 57121 | 2 | 1 | 9 | 2 |
| 19 | 35213 | 15300 | 17261 | 13519 | 28384 | 11721 | 72476 | 37629 | 26847 | 10675 | 16143 | 52342 | 12562 | 17301 | 25979 | 23873 | 11366 | 4620 | 4001 | 3946 | 80520 |
| 91 | 50 | 50 | 70 | 30 | 70 | 00 | 4 | 9 | 4 | 700 | 1 | 1 | 20 | 1 | 5 | 0 | 3 | 6 | 0 | 7 | 9 |
| 19 | 68627 | 30268 | 13005 | 14489 | 11194 | 23231 | 95134 | 58519 | 30293 | 21579 | 85738 | 12959 | 42010 | 10081 | 13883 | 20846 | 19155 | 9120 | 3707 | 3210 | 67774 |
| 92 | 20 | 30 | 20 | 10 | 50 | 90 | 9 |  | 4 | 2 | 90 | 5 | 4 | 30 | 4 | 7 | 9 | 2 | 5 | 4 | 4 |
| 19 | 71137 | 58948 | 25556 | 10771 | 11753 | 89202 | 18276 | 74246 | 45463 | 23478 | 16703 | 66325 | 10021 | 32481 | 77938 | 10732 | 16114 | 1480 | 7049 | 2865 | 54868 |
| 93 | 60 | 00 | 70 | 50 | 00 | 1 | 90 | 5 | 6 | 3 | 8 | 60 | 8 | 4 | 3 | 6 | 8 | 75 | 8 | 8 | 6 |
| 19 | 72401 | 61063 | 49458 | 20900 | 85685 | 91305 | 68130 | 13811 | 55768 | 34040 | 17550 | 12475 | 49514 |  | 24239 | 58156 |  | 1202 | 1104 | 5259 | 43074 |
| 94 | 70 | 50 | 20 | 80 | 7 | 3 | 0 | 40 | 9 | 0 | 0 | 5 | 10 | 74798 |  | 4 | 80081 | 36 | 80 | 8 | 2 |
| 19 | 48254 | 62140 | 51168 | 40345 | 16561 | 66227 | 69322 | 51151 | 10304 | 41467 | 25266 | 13015 |  | 36693 |  | 17958 | 43085 | 5932 | 8907 | 8184 | 35805 |
| 95 | 80 | 20 | 10 | 30 | 20 | 6 | 8 | 8 | 10 | 1 | 8 | 2 | 92476 | 40 | 55422 | 7 | 7 | 7 | 3 | 4 | 3 |
| 19 | 24417 | 41356 | 51370 | 40624 | 30660 | 12123 | 47196 | 48577 | 35504 | 71159 | 28562 | 17380 |  |  | 25207 |  | 12334 | 2959 | 4074 | 6117 | 30209 |
| 96 | 30 | 30 | 40 | 30 | 30 | 40 | 3 | 4 | 1 | 2 | 4 | 3 | 89465 | 63542 | 50 | 38069 | 6 | 10 | 4 | 2 | 0 |
| 19 | 16864 | 20936 | 34347 | 41166 | 31318 | 22867 | 88296 | 33865 | 34563 | 25149 | 50289 | 20161 | 12260 |  |  | 17770 |  | 8694 | 2085 | 2871 | 25604 |
| 97 | 90 | 90 | 90 | 30 | 60 | 30 | 5 | 5 | 8 | 1 | 1 | 5 | 7 | 63090 | 44801 | 70 | 26836 | 6 | 80 | 9 | 0 |
| 19 | 28822 | 14437 | 17127 | 26701 | 30284 | 21977 | 15512 | 58640 | 22222 | 22536 | 16344 | 32627 | 13069 |  |  |  | 11507 | 1737 | 5629 | 1350 | 18436 |
| 98 | 40 | 80 | 10 | 60 | 20 | 90 | 80 | 8 | 5 | 5 | 2 | 2 | 0 | 79437 | 40865 | 29014 | 50 | 6 | 6 | 48 | 2 |
| 19 | 32065 | 24716 | 12001 | 13749 | 20642 | 22668 | 16073 | 11182 | 41923 | 15818 | 16005 | 11594 | 23131 |  |  |  |  | 8152 | 1231 | 3988 | 22626 |
| 99 | 90 | 50 | 80 | 60 | 00 | 00 | 90 | 00 | 5 | 2 | 7 | 4 | 3 | 92623 | 56289 | 28953 | 20555 | 24 | 0 | 0 | 2 |
| 20 | 23860 | 27497 | 20541 | 96303 | 10621 | 15435 | 16559 | 11572 | 79838 | 29801 | 11219 | 11338 |  | 16371 |  |  |  | 1454 | 5767 |  | 18827 |
| 00 | 90 | 20 | 30 | 2 | 20 | 50 | 60 | 00 | 2 | 7 | 1 | 9 | 82088 | 4 | 65542 | 39827 | 20484 | 2 | 23 | 8708 | 0 |
| 20 | 16647 | 20481 | 23064 | 16789 | 76540 | 82418 | 11773 | 12495 | 86789 | 59684 | 22241 |  |  |  | 12197 |  |  | 1526 | 1083 | 4296 | 14672 |
| 01 | 700 | 20 | 00 | 60 | 6 | 8 | 60 | 60 | 2 | 6 | 7 | 83659 | 84514 | 61169 | 7 | 48829 | 29669 | 0 | 3 | 06 | 7 |
| 20 | 19728 | 14281 | 17082 | 18638 | 13112 | 58056 | 61218 | 86309 | 90923 | 62902 | 43170 | 16070 |  |  |  |  |  | 2141 | 1101 |  | 41589 |
| 02 | 70 | 100 | 00 | 70 | 50 | 8 | 5 | 1 | 3 | 6 | 0 | 7 | 60414 | 61013 | 44152 | 88035 | 35239 | 1 | 2 | 7817 | 1 |
| 20 | 17376 | 16931 | 11958 | 13915 | 14736 | 10106 | 43935 | 45798 | 64151 | 67348 | 46510 | 31890 | 11866 |  |  |  |  | 2600 | 1580 |  | 31264 |
| 03 | 30 | 30 | 500 | 00 | 90 | 40 | 0 | 9 | 4 | 4 | 3 | 9 | 1 | 44597 | 45032 | 32585 | 64967 | 5 | 0 | 8126 | 9 |
| 20 | 24497 | 14915 | 14204 | 97778 | 11065 | 11443 | 77152 | 33182 | 34381 | 48004 | 50313 | 34716 | 23793 |  |  |  |  | 4844 | 1939 | 1178 | 23918 |
| 04 | 40 | 40 | 10 | 70 | 60 | 80 |  |  |  | 7 | 7 | 7 | 6 | 88511 | 33261 | 33583 | 24299 | 5 | 1 | 1 | 2 |


| 20 | 15009 | 21041 | 12590 | 11758 | 79247 | 88078 | 89914 | 60128 | 25742 | 26607 | 37102 | 38862 | 26805 | 18368 |  |  |  | 1875 | 3739 | 1496 | 19368 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 05 | 30 | 60 | 10 | 00 | 50 | 2 | 3 | 0 | 1 | 0 | 7 | 1 | 9 | 5 | 68323 | 25673 | 25920 | 4 | 0 | 6 | 7 |
| 20 | 13013 | 12890 | 17741 | 10398 | 94969 | 62797 | 68844 | 69678 | 46369 | 19800 | 20438 | 28481 | 29821 | 20565 | 14091 |  |  | 1988 | 1438 | 2867 | 16003 |
| 06 | 30 | 50 | 60 | 80 | 3 | 20 | 4 | 1 | 3 | 5 | 6 | 6 | 4 | 9 | 0 | 52409 | 19692 | 1 | 5 | 8 | 4 |
| 20 | 23863 | 11180 | 10906 | 14756 | 84900 | 76316 | 49891 | 54306 | 54742 | 36352 | 15505 | 15996 | 22284 | 23329 | 16087 | 11021 |  | 1540 | 1555 | 1125 | 14759 |
| 07 | 20 | 40 | 80 | 30 | 1 | 2 | 60 | 8 | 8 | 1 | 9 | 5 | 7 | 2 | 1 | 7 | 40992 | 2 | 0 | 0 | 3 |
| 20 | 63901 | 20509 | 94937 | 91366 | 12181 | 69210 | 61655 | 40079 | 43487 | 43762 | 29035 | 12379 | 12767 | 17784 | 18617 | 12837 |  | 3270 | 1229 | 1240 | 12674 |
| 08 | 90 | 90 | 0 | 6 | 30 | 7 | 1 | 90 | 6 | 3 | 1 | 3 | 9 | 7 | 0 | 2 | 87948 | 9 | 0 | 8 | 3 |
| 20 | 13691 | 54906 | 17369 | 79108 | 74808 | 98251 | 55225 | 48865 | 31644 | 34265 | 34445 | 22841 |  | 10040 | 13983 | 14637 | 10092 | 6914 | 2571 |  | 10939 |
| 09 | 00 | 70 | 70 | 5 | 6 | 4 | 5 | 5 | 60 | 4 | 9 | 7 | 97359 | 0 | 7 | 4 | 7 | 4 | 5 | 9662 | , |
| 20 | 97218 | 11756 | 46235 | 14309 | 63642 | 58973 | 76333 | 42515 | 37425 | 24169 | 26134 | 26253 | 17402 |  |  | 10649 | 11146 | 7685 | 5265 | 1958 |  |
| 10 | 8 | 60 | 80 | 50 | 2 | 6 | 8 | 8 | 5 | 80 | 6 | 2 | 3 | 74159 | 76466 | 5 | 8 | 7 | 3 | 2 | 90659 |
| 20 | 44158 | 83447 | 98602 | 37783 | 11369 | 49365 | 44962 | 57571 | 31870 | 27963 | 18028 | 19477 | 19557 | 12960 |  |  |  | 8299 | 5722 | 3920 |  |
| 11 | 5 | 8 | 3 | 90 | 60 | 8 | 0 | 7 | 1 | 3 | 80 | 7 | 2 | 5 | 55223 | 56937 | 79292 | 1 | 1 | 0 | 82072 |
| 20 | 27999 | 37901 | 69947 | 80485 | 29968 | 87988 | 37536 | 33811 | 43024 | 23738 | 20792 | 13393 | 14463 | 14518 |  |  |  | 5884 | 6159 | 4246 |  |
| 12 | 10 | 2 | 1 | 5 | 10 | 7 | 3 | 9 | 4 | 1 | 6 | 90 | 6 | 9 | 96203 | 40987 | 42257 | 6 | 1 | 5 | 89997 |
| 20 | 98087 | 24035 | 31816 | 57264 | 64127 | 23329 | 67367 | 28440 | 25467 | 32305 | 17795 | 15574 | 10028 | 10826 | 10866 |  |  | 3162 | 4403 | 4608 |  |
| 13 | 3 | 30 | 2 | 0 | 8 | 90 | 0 | 6 | 7 | 2 | 2 | 2 | 00 | 3 | 2 | 71994 | 30672 | 1 | 3 | 6 | 99115 |
| 20 | 36009 | 84187 | 20145 | 25967 | 45411 | 49617 | 17735 | 50647 | 21248 | 18964 | 24014 | 13216 | 11561 | 74424 |  |  |  | 2275 | 2346 | 3266 | 10772 |
| 14 | 50 | 7 | 70 | 7 | 4 | 9 | 00 | 3 | 3 | 0 | 2 | 5 | 6 | 3 | 80337 | 80627 | 53417 | 6 | 0 | 9 | 3 |
| 20 | 28710 | 30914 | 70734 | 16522 | 20747 | 35479 | 38149 | 13499 | 38331 | 16032 | 14287 | 18077 |  |  | 55980 |  |  | 4017 | 1711 | 1764 | 10557 |
| 15 | 80 | 60 | 9 | 50 | 4 | 8 | 0 | 10 | 4 | 8 | 0 | 2 | 99448 | 86975 | 5 | 60424 | 60639 | 3 | 4 | 3 | 8 |
| 20 | 36749 | 24663 | 26118 | 58657 | 13427 | 16572 | 27992 | 29865 | 10521 | 29807 | 12452 | 11089 | 14027 |  |  | 43423 |  | 4703 | 3115 | 1327 |  |
| 16 | 10 | 00 | 30 | 3 | 70 | 9 | 4 | 5 | 60 | 1 | 4 | 6 | 0 | 77153 | 67470 | 5 | 46868 | 4 | 9 | 4 | 95571 |
| 20 | 50847 | 31565 | 20822 | 21630 | 47573 | 10697 | 13035 | 21840 | 23196 | 81524 | 23066 |  |  | 10842 |  |  | 33558 | 3621 | 3634 | 2407 |  |
| 17 | 00 | 80 | 90 | 20 | 7 | 70 | 3 | 8 | 6 | 6 | 7 | 96303 | 85735 | 4 | 59631 | 52144 | 1 | 9 | 7 | 9 | 8110 |

Table 7.3.1.2. Western horse mackerel. Final assessment. Fishing mortality-at-age.

| year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0.0005 | 0.0046 | 0.0093 | 0.0143 | 0.0186 | 0.0217 | 0.0236 | 0.0247 | 0.0253 | 0.0256 | 0.0257 | 0.0258 | 0.0259 | 0.0259 | 0.0259 | 0.0259 | 0.0259 | 0.0259 | 0.0259 | 0.0259 | 0.0259 |
| 1983 | 0.0006 | 0.0061 | 0.0122 | 0.0188 | 0.0245 | 0.0285 | 0.0311 | 0.0325 | 0.0333 | 0.0337 | 0.0339 | 0.0340 | 0.0340 | 0.0341 | 0.0341 | 0.0341 | 0.0341 | 0.0341 | 0.0341 | 0.0341 | 0.0341 |
| 1984 | 0.0006 | 0.0055 | 0.0109 | 0.0169 | 0.0219 | 0.0256 | 0.0279 | 0.0292 | 0.0298 | 0.0302 | 0.0304 | 0.0305 | 0.0305 | 0.0306 | 0.0306 | 0.0306 | 0.0306 | 0.0306 | 0.0306 | 0.0306 | 0.0306 |
| 1985 | 0.0005 | 0.0045 | 0.0091 | 0.0140 | 0.0182 | 0.0212 | 0.0231 | 0.0241 | 0.0247 | 0.0250 | 0.0252 | 0.0252 | 0.0253 | 0.0253 | 0.0253 | 0.0253 | 0.0253 | 0.0253 | 0.0253 | 0.0253 | 0.0253 |
| 1986 | 0.0006 | 0.0056 | 0.0112 | 0.0173 | 0.0225 | 0.0262 | 0.0285 | 0.0298 | 0.0305 | 0.0309 | 0.0311 | 0.0312 | 0.0312 | 0.0313 | 0.0313 | 0.0313 | 0.0313 | 0.0313 | 0.0313 | 0.0313 | 0.0313 |
| 1987 | 0.0007 | 0.0071 | 0.0142 | 0.0218 | 0.0284 | 0.0331 | 0.0361 | 0.0377 | 0.0386 | 0.0391 | 0.0393 | 0.0395 | 0.0395 | 0.0396 | 0.0396 | 0.0396 | 0.0396 | 0.0396 | 0.0396 | 0.0396 | 0.0396 |
| 1988 | 0.0008 | 0.0078 | 0.0157 | 0.0242 | 0.0315 | 0.0367 | 0.0400 | 0.0418 | 0.0428 | 0.0433 | 0.0436 | 0.0437 | 0.0438 | 0.0439 | 0.0439 | 0.0439 | 0.0439 | 0.0439 | 0.0439 | 0.0439 | 0.0439 |
| 1989 | 0.0008 | 0.0080 | 0.0160 | 0.0246 | 0.0320 | 0.0373 | 0.0407 | 0.0425 | 0.0435 | 0.0441 | 0.0443 | 0.0445 | 0.0445 | 0.0446 | 0.0446 | 0.0446 | 0.0446 | 0.0446 | 0.0446 | 0.0447 | 0.0447 |
| 1990 | 0.0012 | 0.0111 | 0.0222 | 0.0342 | 0.0445 | 0.0519 | 0.0565 | 0.0592 | 0.0605 | 0.0613 | 0.0616 | 0.0618 | 0.0619 | 0.0620 | 0.0620 | 0.0620 | 0.0621 | 0.0621 | 0.0621 | 0.0621 | 0.0621 |
| 1991 | 0.0013 | 0.0125 | 0.0251 | 0.0387 | 0.0503 | 0.0587 | 0.0639 | 0.0669 | 0.0684 | 0.0692 | 0.0697 | 0.0699 | 0.0700 | 0.0701 | 0.0701 | 0.0701 | 0.0702 | 0.0702 | 0.0702 | 0.0702 | 0.0702 |
| 1992 | 0.0020 | 0.0192 | 0.0384 | 0.0593 | 0.0771 | 0.0899 | 0.0979 | 0.1024 | 0.1049 | 0.1061 | 0.1067 | 0.1071 | 0.1072 | 0.1074 | 0.1074 | 0.1075 | 0.1075 | 0.1075 | 0.1075 | 0.1075 | 0.1075 |
| 1993 | 0.0027 | 0.0255 | 0.0511 | 0.0788 | 0.1025 | 0.1195 | 0.1301 | 0.1362 | 0.1394 | 0.1410 | 0.1419 | 0.1423 | 0.1426 | 0.1427 | 0.1428 | 0.1428 | 0.1429 | 0.1429 | 0.1429 | 0.1429 | 0.1429 |
| 1994 | 0.0028 | 0.0268 | 0.0537 | 0.0827 | 0.1076 | 0.1254 | 0.1366 | 0.1430 | 0.1463 | 0.1480 | 0.1489 | 0.1494 | 0.1497 | 0.1498 | 0.1499 | 0.1499 | 0.1500 | 0.1500 | 0.1500 | 0.1500 | 0.1500 |
| 1995 | 0.0043 | 0.0403 | 0.0807 | 0.1245 | 0.1619 | 0.1888 | 0.2056 | 0.2151 | 0.2202 | 0.2228 | 0.2242 | 0.2249 | 0.2252 | 0.2255 | 0.2256 | 0.2257 | 0.2257 | 0.2258 | 0.2258 | 0.2258 | 0.2258 |
| 1996 | 0.0038 | 0.0357 | 0.0714 | 0.1102 | 0.1433 | 0.1670 | 0.1819 | 0.1904 | 0.1948 | 0.1971 | 0.1983 | 0.1989 | 0.1993 | 0.1995 | 0.1996 | 0.1997 | 0.1997 | 0.1997 | 0.1998 | 0.1998 | 0.1998 |
| 1997 | 0.0054 | 0.0509 | 0.1018 | 0.1570 | 0.2042 | 0.2380 | 0.2593 | 0.2713 | 0.2777 | 0.2810 | 0.2826 | 0.2835 | 0.2840 | 0.2843 | 0.2845 | 0.2846 | 0.2846 | 0.2847 | 0.2847 | 0.2847 | 0.2847 |
| 1998 | 0.0037 | 0.0348 | 0.0697 | 0.1074 | 0.1397 | 0.1628 | 0.1774 | 0.1856 | 0.1899 | 0.1922 | 0.1933 | 0.1940 | 0.1943 | 0.1945 | 0.1946 | 0.1947 | 0.1947 | 0.1947 | 0.1948 | 0.1948 | 0.1948 |
| 1999 | 0.0037 | 0.0350 | 0.0701 | 0.1082 | 0.1407 | 0.1640 | 0.1786 | 0.1869 | 0.1913 | 0.1935 | 0.1947 | 0.1953 | 0.1957 | 0.1958 | 0.1960 | 0.1960 | 0.1961 | 0.1961 | 0.1961 | 0.1962 | 0.1962 |
| 2000 | 0.0027 | 0.0258 | 0.0517 | 0.0797 | 0.1036 | 0.1208 | 0.1316 | 0.1377 | 0.1409 | 0.1426 | 0.1435 | 0.1439 | 0.1441 | 0.1443 | 0.1444 | 0.1444 | 0.1445 | 0.1445 | 0.1445 | 0.1445 | 0.1445 |
| 2001 | 0.0033 | 0.0315 | 0.0630 | 0.0972 | 0.1264 | 0.1474 | 0.1605 | 0.1679 | 0.1719 | 0.1739 | 0.1750 | 0.1755 | 0.1758 | 0.1760 | 0.1761 | 0.1762 | 0.1762 | 0.1762 | 0.1762 | 0.1763 | 0.1763 |
| 2002 | 0.0029 | 0.0275 | 0.0551 | 0.0849 | 0.1104 | 0.1287 | 0.1402 | 0.1467 | 0.1501 | 0.1519 | 0.1528 | 0.1533 | 0.1536 | 0.1537 | 0.1538 | 0.1539 | 0.1539 | 0.1539 | 0.1539 | 0.1540 | 0.1540 |
| 2003 | 0.0027 | 0.0256 | 0.0513 | 0.0791 | 0.1029 | 0.1200 | 0.1307 | 0.1367 | 0.1399 | 0.1416 | 0.1425 | 0.1429 | 0.1431 | 0.1433 | 0.1434 | 0.1434 | 0.1435 | 0.1435 | 0.1435 | 0.1435 | 0.1435 |
| 2004 | 0.0021 | 0.0195 | 0.0390 | 0.0601 | 0.0782 | 0.0912 | 0.0993 | 0.1039 | 0.1064 | 0.1076 | 0.1083 | 0.1086 | 0.1088 | 0.1089 | 0.1090 | 0.1090 | 0.1090 | 0.1090 | 0.1090 | 0.1091 | 0.1091 |
| 2005 | 0.0022 | 0.0206 | 0.0412 | 0.0636 | 0.0827 | 0.0964 | 0.1050 | 0.1098 | 0.1124 | 0.1137 | 0.1144 | 0.1148 | 0.1150 | 0.1151 | 0.1152 | 0.1152 | 0.1152 | 0.1153 | 0.1153 | 0.1153 | 0.1153 |
| 2006 | 0.0018 | 0.0171 | 0.0342 | 0.0528 | 0.0687 | 0.0801 | 0.0872 | 0.0912 | 0.0934 | 0.0945 | 0.0951 | 0.0954 | 0.0955 | 0.0956 | 0.0957 | 0.0957 | 0.0957 | 0.0957 | 0.0957 | 0.0958 | 0.0958 |
| 2007 | 0.0014 | 0.0135 | 0.0271 | 0.0418 | 0.0543 | 0.0633 | 0.0690 | 0.0722 | 0.0739 | 0.0747 | 0.0752 | 0.0754 | 0.0756 | 0.0756 | 0.0757 | 0.0757 | 0.0757 | 0.0757 | 0.0757 | 0.0758 | 0.0758 |
| 2008 | 0.0017 | 0.0162 | 0.0324 | 0.0499 | 0.0650 | 0.0757 | 0.0825 | 0.0863 | 0.0883 | 0.0894 | 0.0899 | 0.0902 | 0.0904 | 0.0904 | 0.0905 | 0.0905 | 0.0906 | 0.0906 | 0.0906 | 0.0906 | 0.0906 |
| 2009 | 0.0023 | 0.0219 | 0.0438 | 0.0675 | 0.0878 | 0.1024 | 0.1115 | 0.1167 | 0.1195 | 0.1209 | 0.1216 | 0.1220 | 0.1222 | 0.1223 | 0.1224 | 0.1224 | 0.1225 | 0.1225 | 0.1225 | 0.1225 | 0.1225 |


| 2010 | 0.0027 | 0.0259 | 0.0519 | 0.0800 | 0.1040 | 0.1213 | 0.1321 | 0.1382 | 0.1415 | 0.1431 | 0.1440 | 0.1444 | 0.1447 | 0.1448 | 0.1449 | 0.1450 | 0.1450 | 0.1450 | 0.1450 | 0.1451 | 0.1451 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2011 | 0.0028 | 0.0265 | 0.0530 | 0.0817 | 0.1063 | 0.1240 | 0.1350 | 0.1413 | 0.1446 | 0.1463 | 0.1472 | 0.1476 | 0.1479 | 0.1480 | 0.1481 | 0.1482 | 0.1482 | 0.1482 | 0.1482 | 0.1483 | 0.1483 |
| 2012 | 0.0026 | 0.0250 | 0.0501 | 0.0772 | 0.1004 | 0.1171 | 0.1275 | 0.1334 | 0.1365 | 0.1382 | 0.1390 | 0.1394 | 0.1397 | 0.1398 | 0.1399 | 0.1399 | 0.1400 | 0.1400 | 0.1400 | 0.1400 | 0.1400 |
| 2013 | 0.0028 | 0.0265 | 0.0531 | 0.0819 | 0.1065 | 0.1242 | 0.1353 | 0.1415 | 0.1449 | 0.1466 | 0.1475 | 0.1479 | 0.1482 | 0.1483 | 0.1484 | 0.1485 | 0.1485 | 0.1485 | 0.1485 | 0.1486 | 0.1486 |
| 2014 | 0.0026 | 0.0241 | 0.0483 | 0.0744 | 0.0968 | 0.1129 | 0.1229 | 0.1286 | 0.1316 | 0.1332 | 0.1340 | 0.1344 | 0.1346 | 0.1348 | 0.1349 | 0.1349 | 0.1349 | 0.1350 | 0.1350 | 0.1350 | 0.1350 |
| 2015 | 0.0020 | 0.0186 | 0.0372 | 0.0574 | 0.0747 | 0.0870 | 0.0948 | 0.0992 | 0.1015 | 0.1027 | 0.1033 | 0.1037 | 0.1038 | 0.1039 | 0.1040 | 0.1040 | 0.1041 | 0.1041 | 0.1041 | 0.1041 | 0.1041 |
| 2016 | 0.0020 | 0.0193 | 0.0385 | 0.0594 | 0.0773 | 0.0901 | 0.0982 | 0.1027 | 0.1051 | 0.1064 | 0.1070 | 0.1073 | 0.1075 | 0.1076 | 0.1077 | 0.1077 | 0.1077 | 0.1078 | 0.1078 | 0.1078 | 0.1078 |
| 2017 | 0.0017 | 0.0160 | 0.0321 | 0.0495 | 0.0643 | 0.0750 | 0.0817 | 0.0855 | 0.0875 | 0.0885 | 0.0890 | 0.0893 | 0.0895 | 0.0896 | 0.0896 | 0.0896 | 0.0897 | 0.0897 | 0.0897 | 0.0897 | 0.0897 |

Table 7.3.1.3. Western horse mackerel. Final assessment. Stock summary table.


| Year | Recruit <br> (THOUSA <br> NDS) | Total BIOMAS S | Spawnin <br> G <br> BIOMASS | Catch | $\begin{gathered} \text { YIELD/SS } \\ \text { B } \end{gathered}$ | Fbar(13) | Fbar(4- <br> 8) | $\begin{gathered} \text { Fbar(1- } \\ 10) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | 980873 | 130891 | 1200660 | 165258.35 | 0.1376395 | 0.053855 | 0.130480 | 0.110801 |
|  |  | 0 |  |  | 9 |  | 4 | 2 |
| 2014 | 3600950 | 118886 | 1067470 | 136359.64 | 0.1277409 | 0.0489376 | 0.118565 | 0.100683 |
|  |  | 0 |  |  | 6 | 7 | 4 | 5 |
| 2015 | 2871080 | 111376 | 949935 | 98419.2 | 0.1036062 | 0.0377403 | 0.091437 | 0.077646 |
|  |  | 0 |  |  | 5 | 3 | 6 | 9 |
| 2016 | 3674910 | 110683 | 902625 | 98810 | 0.1094696 | 0.0390763 | 0.094674 | 0.080395 |
|  |  | 0 |  |  |  | 3 | 4 | 5 |
| 2017 | 5084700 | 112729 | 872011 | 82961 | 0.0951375 | 0.0325146 | 0.078777 | 0.066895 |
|  |  | 0 |  |  | 6 | 7 |  | 8 |

Table 7.4.1. Western Horse Mackerel. Short-term prediction: INPUT DATA. *geometric mean of the recruitment time-series from 1983 to 2017.

| Age | N | MAT | M | PF | PM | SwT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2584327* | 0.000 | 0.15 | 0 | 0 | 0.000396 |
| 1 | 3156580 | 0.000 | 0.15 | 0 | 0 | 0.014395 |
| 2 | 2082290 | 0.047 | 0.15 | 0 | 0 | 0.040471 |
| 3 | 2163020 | 0.269 | 0.15 | 0 | 0 | 0.068313 |
| 4 | 475737 | 0.731 | 0.15 | 0 | 0 | 0.099264 |
| 5 | 1069770 | 0.953 | 0.15 | 0 | 0 | 0.130931 |
| 6 | 130353 | 0.993 | 0.15 | 0 | 0 | 0.161589 |
| 7 | 218408 | 0.999 | 0.15 | 0 | 0 | 0.190145 |
| 8 | 231966 | 1.000 | 0.15 | 0 | 0 | 0.216006 |
| 9 | 815246 | 1.000 | 0.15 | 0 | 0 | 0.23894 |
| 10 | 230667 | 1.000 | 0.15 | 0 | 0 | 0.258956 |
| 11 | 96303 | 1.000 | 0.15 | 0 | 0 | 0.27621 |
| 12 | 85735 | 1.000 | 0.15 | 0 | 0 | 0.290939 |
| 13 | 108424 | 1.000 | 0.15 | 0 | 0 | 0.303419 |
| 14 | 59631 | 1.000 | 0.15 | 0 | 0 | 0.313927 |
| 15 | 52144 | 1.000 | 0.15 | 0 | 0 | 0.322733 |
| 16 | 335581 | 1.000 | 0.15 | 0 | 0 | 0.330083 |
| 17 | 36219 | 1.000 | 0.15 | 0 | 0 | 0.336199 |
| 18 | 36347 | 1.000 | 0.15 | 0 | 0 | 0.341275 |
| 19 | 24079 | 1.000 | 0.15 | 0 | 0 | 0.345479 |
| 20 | 84110 | 1.000 | 0.15 | $0$ | 0 | 0.352296 |

Table 7.4.2. Western Horse Mackerel. Short-term prediction; single area management option table. OPTION: Catch constraint 115,470 tin 2018 (EU TAC).

| $\begin{gathered} \text { Scenari } \\ \text { os } \end{gathered}$ | Ffacto <br> R | Fbar | $\begin{gathered} \text { CATCH_20 } \\ 17 \end{gathered}$ | $\begin{gathered} \text { CATCH_20 } \\ 18 \end{gathered}$ | $\begin{gathered} \text { CATCH_20 } \\ 19 \end{gathered}$ | $\begin{gathered} \text { CATCH_20 } \\ 20 \end{gathered}$ | $\begin{gathered} \text { SSB_201 } \\ 9 \end{gathered}$ | $\begin{gathered} \text { SSB_202 } \\ 0 \end{gathered}$ | $\begin{gathered} \text { CHANGE_SSB_2019- } \\ 2020(\%) \end{gathered}$ | $\begin{gathered} \text { CHANGE_CATCH_2018- } \\ 2019(\%) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{F}_{0}$ | 0.00 | $\begin{gathered} 0.00 \\ 0 \end{gathered}$ | 82961 | 115470 | 0 | 0 | 941821 | 1113644 | 18.24 | -100.000 |
|  | 0.10 | $\begin{gathered} 0.00 \\ 7 \end{gathered}$ | 82961 | 115470 | 9506 | 10904 | 941821 | 1105377 | 17.37 | -91.767 |
|  | 0.20 | $\begin{gathered} 0.01 \\ 3 \end{gathered}$ | 82961 | 115470 | 18944 | 21581 | 941821 | 1097174 | 16.49 | -83.594 |
|  | 0.30 | $\begin{gathered} 0.02 \\ 0 \end{gathered}$ | 82961 | 115470 | 28313 | 32034 | 941821 | 1089035 | 15.63 | -75.480 |
|  | 0.40 | $\begin{gathered} 0.02 \\ 7 \end{gathered}$ | 82961 | 115470 | 37614 | 42268 | 941821 | 1080960 | 14.77 | -67.425 |
|  | 0.50 | $\begin{gathered} 0.03 \\ 3 \end{gathered}$ | 82961 | 115470 | 46848 | 52287 | 941821 | 1072947 | 13.92 | -59.428 |
|  | 0.60 | $\begin{gathered} 0.04 \\ 0 \end{gathered}$ | 82961 | 115470 | 56016 | 62095 | 941821 | 1064998 | 13.08 | -51.489 |
|  | 0.70 | $\begin{gathered} 0.04 \\ 7 \end{gathered}$ | 82961 | 115470 | 65117 | 71694 | 941821 | 1057110 | 12.24 | -43.607 |
|  | 0.80 | $\begin{gathered} 0.05 \\ 4 \\ \hline \end{gathered}$ | 82961 | 115470 | 74152 | 81090 | 941821 | 1049284 | 11.41 | -35.782 |
|  | 0.90 | $\begin{gathered} 0.06 \\ 0 \end{gathered}$ | 82961 | 115470 | 83123 | 90285 | 941821 | 1041518 | 10.59 | -28.014 |
| $\mathrm{F}_{\text {stq }}$ | 1.00 | $\begin{gathered} 0.06 \\ 7 \end{gathered}$ | 82961 | 115470 | 92028 | 99283 | 941821 | 1033814 | 9.77 | -20.301 |
|  | 1.10 | $\begin{gathered} 0.07 \\ 4 \\ \hline \end{gathered}$ | 82961 | 115470 | 100870 | 108088 | 941821 | 1026169 | 8.96 | -12.644 |
|  | 1.20 | $\begin{gathered} 0.08 \\ 0 \end{gathered}$ | 82961 | 115470 | 109648 | 116702 | 941821 | 1018584 | 8.15 | -5.042 |
|  | 1.30 | $\begin{gathered} 0.08 \\ 7 \end{gathered}$ | 82961 | 115470 | 118363 | 125131 | 941821 | 1011058 | 7.35 | 2.505 |
|  | 1.40 | $\begin{gathered} 0.09 \\ 4 \end{gathered}$ | 82961 | 115470 | 127015 | 133376 | 941821 | 1003591 | 6.56 | 9.998 |



### 7.15 Figures



Figure 7.1.1.1: Western horse mackerel. Catch by quarter and year for 2000-2017


Figure 7.1.2.1. Western horse mackerel. Catch categories since 2000.


Figure 7.1.3.1: Western horse mackerel. Catch by ICES Division and year for 1982-2017


Figure 7.2.1.1: Western horse mackerel. annual egg production curve for western horse mackerel. The curves for 2004, 20072010 and 2013 are included for comparison.


Figure 7.2.2.1: Western horse mackerel. Trend of the fisheries independent indices of abundance used in the assessment of Western Horse mackerel -- Plot on top: Spawning index from egg survey; plot in the middle: recruitment index from IBTS survey; plot at the bottom: biomass estimates from Pelacus acoustic survey. Confidence intervals are shown as well.


Figure 7.2.2.2: Western horse mackerel. spatial distribution estimated during PELACUS 0318 in the Cantabrian Sea (NASC values and density -biomass expressed as tonnes per square nautical mile)


Figure 7.2.2.3: Western horse mackerel abundance and biomass estimates by age group during PELACUS 0318.


Figure 7.2.2.4: Western horse mackerel. Spatial distribution estimated during PELGAS 2018 (NASC values)

Horse mackerel


Figure 7.2.2.5: Western horse mackerel biomass estimates (2000-18) during PELGAS

2017 Western Stock: cat@ge by division


Figure 7.2.4.1: Western horse mackerel.. Catch-at-age matrix by division in 2017, expressed as numbers (millions)


Figure 7.2.4.2: Western horse mackerel.. Catch-at-age matrix by year, expressed as numbers (millions).


Figure 7.2.4.1: Western horse mackerel. Catch-at-age matrix, expressed as numbers (thousands). The area of bubbles is proportional to the catch number. Note that age 15 is a plus group.


Figure 7.2.5.1: Western horse mackerel. Weight at age in the catch (kg) by year.


Figure 7.2.5.2: Western horse mackerel. Weight at length in the stock (kg) as estimated by SS.


Figure 7.2.6.1: Western horse mackerel. Maturity-at-age as used in the assessment model.


Figure 7.2.10.1: Western horse mackerel. Length frequency distribution of the catch data as used in the assessment model.


Figure 7.2.10.2: Western horse mackerel. Within-cohort consistency in the catch-at-age matrix, shown by plotting the log-catch of a cohort at a particular age against the log-catch of the same cohort at subsequent ages. Thick lines represent a significant ( $p<0.05$ ) regression and the curved lines are approximate $95 \%$ confidence intervals.


Figure 7.2.10.3: Western horse mackerel. Catch numbers at age composition by decade.


Figure 7.2.10.4: Western horse mackerel. Data exploration. Correlation plot between indices of abundance.


Figure 7.2.11.1: Western horse mackerel. Model fitting. Fitting of the model to the fisheries independent indices.


Figure 7.2.11.1: Western horse mackerel. Model fitting. Fitting of the model to the catch-at-age matrix from 1982 to 2002.


Figure 7.2.11.1 (cont'd): Western horse mackerel. Model fitting. Fitting of the model to the length compostion of the catch data from 2002 to 2017.


Figure 7.2.11.1 (cont'd): Western horse mackerel. Model fitting. Fitting of the model to the length composition of the acoustic survey.


Figure 7.2.11.1 (cont'd): Western horse mackerel. Model fitting. Fitting of the model to the Age length comp of the catch.


Figure 7.2.11.2: Western horse mackerel. Retrospective analysis. 10 years of retrospective analysis for SSB (left), Recruitment (middle), and F (right).


Figure 7.3.1.1: Western horse mackerel. Model results. Spawning-stock biomass ( 0.5 of the overall SSB only is shown; plot on the left) and recruitment estimates (plot on the right) from the assessment model from 1982 to $2017.95 \%$ CI are shown as well.


Figure 7.2.11.2: Western horse mackerel. Model results. Fishing mortality estimates ( $\mathrm{F}_{\text {bar }}$ ages 1-10) from the assessment model from 1982 to $2017.95 \%$ CI intervals are shown as well.


Figure 7.2.11.3: Western horse mackerel. Model results. Historical assessment results. Note: since the 2017 assessment, SSB is estimated on 1st of January. Prior to 2017 SSB has been estimated in May (spawning time).

## 8 Northeast Atlantic Mackerel

### 8.1 ICES Advice and International Management Applicable to 2017

From 2001 to 2007 the internationally agreed TACs covered most of the distribution area of the Northeast Atlantic mackerel. From 2008 to 2014, no agreement has been reached among the Coastal States on the sharing of the mackerel quotas. In 2014, three of the Coastal States agreed on a Management Strategy for 2015 and the subsequent five years. However, the total declared quotas for 2015 to 2018 all exceed the TAC advised by ICES. An overview of the declared quotas and transfers for 2018, as available to WGWIDE, is given in the text table below. Total removals of mackerel are expected to be approximately 1000559 t in 2018, exceeding the ICES advice for 2018 by about 450000 t , and the agreed TAC by the three Coastal States (EU, NO and FO) and when employing the $-20 \%$ interannual TAC stabiliser in the management rule by about 184000 t.

| Estimation of 2018 catch | Tonnes | Reference |
| :--- | ---: | :--- |
| EU quota | 404815 | European Council Regulation 2018/120 |
| Norwegian quota | 183857 | Directorate of Fisheries in Norway |
| Inter-annual quota transfer 2017->2018 <br> (NO) | -8621 | Directorate of Fisheries in Norway |
| Russian quota | 109415 | NEAFC HOD 18/18 |
| Discards | 2832 | Previous years estimate |
| Icelandic quota | 134772 | Icelandic regulation No. 351/2018 |
| Faroese quota | 102924 | Faroese regulation No. 1/2018 |
| Inter-annual quota transfer 2017->2018 <br> (FO) | 66365 | Ministry of Fisheries, Hunting and <br> Agriculture in Greenland |
| Greenland quota | 1000559 |  |
| Total expected catch (incl. discard) ${ }^{1,2}$ |  |  |
| ${ }^{1}$ No guesstimates of banking from 2018 to 2019 |  |  |
| ${ }^{2}$ Quotas refer to claims by each party for 2018 |  |  |

The quota figures and transfers in the text table above were based on various national regulations, official press releases, and discard estimates.

Various international and national measures to protect mackerel are in operation throughout the mackerel catching countries. Refer to Table 8.2.4.1 for an overview.

Since the mid-1970s, ICES has continuously recommended conservation measures for the North Sea component of the Northeast Atlantic mackerel stock (e.g. ICES, 1974; ICES, 1981). The recommended closure of Division 4. a for fishing during the first half of the year is based on the perception that the western mackerel enter the North Sea in July/August, and remain there until December before migrating to their spawning areas. Updated observations from the late 1990s suggested that this return migration actually started in mid- to late February (Jansen et al., 2012). The EU TAC regulations stated that within the limits of the quota for the western component (ICES Subareas and Divisions $6,7,8 . a, b, d, e, 5 . b(E U)$, 2.a (non-EU), 12, 14), a certain quantity of this stock may be caught in Division 4 .a during the periods 1 January to 15 February and 1 September to 31 December. Up to 2010, $30 \%$ of the Western EU TAC of mackerel (MAC/2CX14-) could be taken in Division 4.a. From 2011 onwards, this percentage has been set at $40 \%$ and from 2015 at $60 \%$.

### 8.2 The Fishery

### 8.2.1 Fleet Composition in 2017

A description of the fleets operated by the major mackerel catching nations is given in Table 8.2.1.

The total fleet can be considered to consist of the following components:
Freezer trawlers. These are commonly large vessels (up to 150 m ) that usually operate a single mid-water pelagic trawl, although smaller vessels may also work as pair trawlers. These vessels are at sea for several weeks and sort and process the catch on board, storing the mackerel in frozen 20 kg blocks. The Dutch, German and the majority of the French and English fleets consist of these vessels which are owned and operated by a small number of Dutch companies. They fish in the North Sea, west of the UK and Ireland and also in the English Channel and further south along the western coast of France. The Russian summer fishery in Division 2.a is also prosecuted by freezer trawlers and partly the Icelandic fishery in Division 5.a and in some years in 14.b.

Purse seiners. The majority of the Norwegian catch is taken by these vessels, targeting mackerel overwintering close to the Norwegian coastline. The largest vessels ( $>20 \mathrm{~m}$ ) used refrigerated seawater (RSW), storing the catch in tanks containing refrigerated seawater (RSW). Smaller purse seiners use ice to chill their catch which they take on prior to departure. A purse seine fleet is also the most important component of the Spanish fleet. They are numerous and target mackerel early in the year close to the northern Spanish coast. These are dry hold vessels, chilling the catch with ice. Denmark also has a purse seine fleet operating in the northern North Sea.
Pelagic trawlers. These vessels vary in size from $20-100 \mathrm{~m}$ and operate both individually and as pairs. The largest of the pelagic trawlers use RSW tanks for storage. Iceland, Greenland, Faroes, Scotland and Ireland fish mackerel using pelagic trawlers. Scottish and Icelandic vessels mostly operate singly whereas Ireland and Faroes vessels tend to use pair trawls. Spain also has a significant trawler fleet which target mackerel with a demersal trawl in Subarea 8 and Division 9.a.N.
Lines and jigging. Norway and England have handline fleets operating inshore in the Skagerrak (Norway) and in Divisions 7.e/f (England) around the coast of Cornwall, where other fishing methods are not permitted. Spain also has a large artisanal handline fleet as do France and Portugal. A small proportion of the total catch reported by Scotland (Divisions 4.a and 4.b) and Iceland (Division 5.a) is taken by a handline fleet.

Gillnets. Gillnet fleets are operated by Norway and Spain.

### 8.2.2 Fleet Behaviour in 2017

The most important changes in recent years are related to the geographical expansion of the northern summer fishery (Subareas 2, 5 and 14) and changes in southern waters due to stricter TAC compliance by Spanish authorities. Fishing in the North Sea and west of the British Isles followed a traditional pattern, targeting mackerel on their spawning migration from the Norwegian deep in the northern North Sea, westwards around the north coast of Scotland and down the west coast of Scotland and Ireland.

The Russian freezer trawler fleet operates over a wide area in northern waters. This fleet targets herring and blue whiting in addition to mackerel. In the third quarter of 2017 the Russian vessels took all their catch in Division 2.a.

Total catches from Icelandic vessels were similar to those in recent years with the majority of the catch taken in Division 5.a in waters south and south-east of Iceland. Catches were also taken to the east and west of Iceland. In 2017, Iceland and Greenland targeted mackerel in Division 14.b, with 3\% of the total catch coming from this area. Catches from Greenland have increased in 2017 to 46 kt from 30 kt in 2016 but are still lower than the 87 kt caught in 2014 which was the biggest catch by this fleet to date.

Concerning the Spanish fisheries, no new regulations have been implemented since 2010 when a new control regime was enforced. Fishery has started as in previous years at the beginning of March, although the southern spawning component was already concentrated at their spawning grounds as earlier as February.

### 8.2.3 Recent Changes in Fishing Technology and Fishing Patterns

Northeast Atlantic mackerel, as a widely distributed species, is targeted by a number of different fishing métiers. Most of the fishing patterns of these métiers have remained unchanged during the most recent years, although the timing of the spawning migration and geographical distribution can change from year to year and this affects the fishery in various areas.

Recent changes are notable for two areas and métiers in particular:
In 2010, the Faroese fleet switched from purse-seining in Norwegian and EU waters to pair trawling in the Faroese area. The Faroese fleet used to catch their mackerel quota in Divisions 4.a and 6.a during September-October with purse-seiners. However, as no agreement has been reached between the Coastal States since 2009, the mackerel quota has been taken in Faroese waters during June-October by the same fleet using pair trawls. The mackerel distribution is more scattered during summer and pair trawls seem to be effective in such circumstances. However, since the agreement between the three of the Coastal States for the fisheries in 2015, parts of the Faroese quota will now again be taken with purse-seines in Divisions 5.a and 6.a. In recent years, up to $25 \%$ of the Faroese quota have been granted to smaller, traditionally demersal trawlers using pair trawls.

Also targeting summer feeding mackerel, Icelandic vessels have increased effort and catch dramatically in recent years from 4 kt in 2006 to an average 160 kt annually since 2011. This fishery operates over a wide area E, NE, SE, S and SW off Iceland. Since 2011, there has been less fishing activity to the north and north-east and an increase in catches taken south and west of Iceland. Greenland has reported catches from Division 14.b since 2011.

In Spain part of the purse seiner fleet is using hand lines instead of nets. Although, neither the number of vessels and its evolution nor the reason for such change were deeply analysed, it seems market reasons are driving this shift.

### 8.2.4 Regulations and their Effects

An overview of the major existing technical measures, effort controls and management plans are given in Table 8.2.4.1. Note that there may be additional existing international and national regulations that are not listed here.

Between 2010 and 2016 no overarching Coastal States Agreement/NEAFC Agreement was in place and no overall international regulation on catch limitation was in force. Currently there is no agreement on a management strategy covering all parties fishing mackerel. In 2014, three of the Coastal States (The EU, Faroes and Norway) agreed on
a Management Strategy for 2015 and the subsequent five years. However, the total declared quotas taken by all parties since 2015 have greatly exceeded the TAC advised by ICES (see Section 8.1).

Management aimed at a fishing mortality in the range of $0.15-0.20$ in the period 1998-2008. The current management plan aims at a fishing mortality in the range $0.20-0.22$. The fishing mortality realised during 1998-2008 was in the range of 0.27 to 0.46 . Implementation of the management plan resulted in a reduced fishing mortality and increased biomass. Since 2008 catches have greatly exceeded those given by the plan.

The measures advised by ICES to protect the North Sea spawning component aim at setting the conditions for making a recovery of this component possible. Before the late 1960s, the North Sea spawning biomass of mackerel was estimated at above 3 million tonnes. The collapse of mackerel in the North Sea in the late 1960s was most likely driven by very high catches and associated fishing mortality. However, the lack of recovery of mackerel in the North Sea was probably associated with unfavourable environmental conditions, particularly reduced temperatures (unfavourable for spawning), lower zoo-plankton availability in the North Sea and increased wind-stress induced turbulence. These unfavourable environmental conditions probably led the mackerel to spawn in western waters instead of in the North Sea.

A review of the mackerel in the North Sea, carried out during WKWIDE 2017 (ICES, 2017a) concluded that Northeast Atlantic mackerel should be considered as a single population (stock) with individuals that show stronger or weaker affinity for spawning in certain parts of the spawning area. Management should ensure that fisheries do not decrease genetic and behavioural diversity, since this could reduce future production. Protection of mackerel that tend to spawn in the north-eastern parts of the spawning area is therefore still advisable to some extent.

In the southern area, a Spanish national regulation affecting mackerel catches of Spanish fisheries has been implemented since 2010. In 2015, fishing opportunity was distributed by region and gear and for the bottom trawl fleet, by individual vessel. This year, Spanish mackerel fishing opportunity in Divisions 8.c and 9.a was established at 39674 t resulting from the quota established (Commission Regulation (EU) No $104 / 2015$. This was reduced by 9797 t due to the scheduling payback quota due to overfishing of the mackerel quota allocated to Spain in 2010 (Commission Regulation No 976/2012).

Within the area of the southwest Mackerel Box off Cornwall in southern England only handliners are permitted to target mackerel. This area was set up at a time of high fishing effort in the area in 1981 by Council Regulation to protect juvenile mackerel, as the area is a well-known nursery. The area of the box was extended to its present size in 1989.

Additionally, there are various other national measures in operation in some of the mackerel catching countries.

The first phase of a landing obligation came into force in 2015 for all EU vessels in pelagic and industrial fisheries. All species that are managed through TACs and quotas must be landed under the obligation unless there is a specific exemption such as de minimis. There are no de minimis exemptions for mackerel.

### 8.3 Quality and Adequacy of sampling Data from Commercial Fishery

The sampling of the commercial catch of North East Atlantic (NEA) mackerel is summarised below:

| Year | WG Total Catch <br> (t) | \% catch covered by sampling programme* | No. Samples | No. <br> Measured | No. <br> Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 760000 | 85 | 920 | 77000 | 11800 |
| 1993 | 825000 | 83 | 890 | 80411 | 12922 |
| 1994 | 822000 | 80 | 807 | 72541 | 13360 |
| 1995 | 755000 | 85 | 1008 | 102383 | 14481 |
| 1996 | 563600 | 79 | 1492 | 171830 | 14130 |
| 1997 | 569600 | 83 | 1067 | 138845 | 16355 |
| 1998 | 666700 | 80 | 1252 | 130011 | 19371 |
| 1999 | 608928 | 86 | 1109 | 116978 | 17432 |
| 2000 | 667158 | 76 | 1182 | 122769 | 15923 |
| 2001 | 677708 | 83 | 1419 | 142517 | 19824 |
| 2002 | 717882 | 87 | 1450 | 184101 | 26146 |
| 2003 | 617330 | 80 | 1212 | 148501 | 19779 |
| 2004 | 611461 | 79 | 1380 | 177812 | 24173 |
| 2005 | 543486 | 83 | 1229 | 164593 | 20217 |
| 2006 | 472652 | 85 | 1604 | 183767 | 23467 |
| 2007 | 579379 | 87 | 1267 | 139789 | 21791 |
| 2008 | 611063 | 88 | 1234 | 141425 | 24350 |
| 2009 | 734889 | 87 | 1231 | 139867 | 28722 |
| 2010 | 869451 | 91 | 1241 | 124695 | 29462 |
| 2011 | 938819 | 88 | 923 | 97818 | 22817 |
| 2012 | 894684 | 89 | 1216 | 135610 | 38365 |
| 2013 | 933165 | 89 | 1092 | 115870 | 25178 |
| 2014 | 1394454 | 90 | 1506 | 117250 | 43475 |
| 2015 | 1208990 | 88 | 2132 | 137871 | 24283 |
| 2016 | 1094066 | 89 | 2200 | 149216 | 21456 |
| 2017 | 1155944 | 87 | 2183 | 151548 | 24104 |

Overall sampling effort in 2017 was similar to previous years with $87 \%$ of the catch sampled. It should be noted that this proportion is based on the total sampled catch. Nations with large, directed fisheries are capable of sampling $100 \%$ of their catch which may conceal deficiencies in sampling elsewhere.

The 2017 sampling levels for countries with a WG catch of greater than 100 t are shown below.

| Country | Offi- <br> cial <br> Catch <br> (t) | \% WG catch covered by sampling programme | No. Samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 128 | 0\% | 0 | 0 | 0 |
| Denmark | 40080 | 93\% | 3 | 214 | 214 |
| Faroe Islands | 99667 | 85\% | 14 | 750 | 712 |
| France | 23800 | 0\% | 0 | 0 | 0 |
| Germany | 24832 | 33\% | 63 | 13562 | 819 |
| Greenland | 46388 | 79\% | 15 | 2395 | 125 |
| Iceland | 167366 | 99\% | 107 | 4209 | 2431 |
| Ireland | 84915 | 98\% | 44 | 7751 | 1587 |
| Netherlands | 43766 | 60\% | 33 | 2174 | 825 |
| Norway | 222356 | 96\% | 73 | 2126 | 2126 |
| Portugal | 634 | 100\% | 136 | 6735 | 766 |
| Russia | 138061 | 98\% | 175 | 54185 | 1503 |
| Spain | 22172 | 100\% | 920 | 15367 | 7295 |
| UK (England \& Wales) | 26463 | 3\% | 74 | 6054 | 3578 |
| UK (Northern Ireland) | 16888 | 0\% | 0 | 0 | 0 |
| UK (Scotland) | 182528 | 97\% | 38 | 4505 | 1081 |

The majority of countries achieved a high level of sampling coverage. Belgian catches are by-catch in the demersal fisheries in the North Sea. France supplied a quantity of length-frequency data to the working group which can be utilised to characterise the selection of the fleet but requires an allocation of catch at age proportions from another sampled fleet in order to raise the data for use in the assessment. England only samples landings from the handline fleet operating off the Cornish coast, representing only a small proportion of the national catch, the remainder reported from freezer trawlers. Cooperation between the Dutch and German sampling programmes (which sampled $60 \%$ and $33 \%$ respectively) is designed to provide complete coverage for the freezer trawlers operating under these national flags and also those of England and France. There is however, an absence of sampling from ICES Division 4.a in quarter 4 for this fleet with landings of 37 kt . Northern Ireland, with a WG catch of 17 kt did not provide any sampling information. Catch sampling levels per ICES Division (for those with a WG catch of $>100 \mathrm{t}$ ) are shown below.

| Division | Official <br> Catch $(\mathbf{t})$ | WG Catch <br> $\mathbf{( t )}$ | No. Sam- <br> ples | No. Measured/ <br> per kt | No. Aged/per <br> $\mathbf{k t}$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
| 2.a | 465355 | 465355 | 287 | $57988 / 126$ | $4714 / 10$ |
| 3.a | 686 | 686 | 0 | $0 / 0$ | $0 / 0$ |
| 4.a | 263825 | 263825 | 87 | $9200 / 35$ | $2338 / 9$ |
| 4.b | 4723 | 4723 | 1 | $87 / 18$ | $25 / 5$ |
| 4.c | 532 | 532 | 0 | $0 / 0$ | $0 / 0$ |
| 5.a | 87734 | 87734 | 66 | $2794 / 32$ | $1510 / 17$ |
| 5.b | 11344 | 11344 | 1 | $165 / 15$ | $122 / 11$ |
| 6.a | 226056 | 226056 | 90 | $16804 / 74$ | $2091 / 9$ |
| 7.b | 6421 | 6421 | 22 | $2306 / 359$ | $435 / 68$ |
| 7.d | 6082 | 6082 | 0 | 0 | 0 |
| 7.e | 956 | 956 | 38 | $2213 / 2314$ | $2074 / 2169$ |
| 7.f | 679 | 679 | 36 | $3841 / 5657$ | $1504 / 2215$ |
| 7.j | 1817 | 1817 | 160 | $366 / 201$ | $3 / 2$ |
| 8.a | 2150 | 2150 | 0 | 0 | 0 |
| 8.b | 4854 | 4854 | 45 | $1866 / 388$ | $4164 / 116$ |
| 8.c | 31059 | 31059 | 362 | $26719 / 860$ | $2106 / 9145$ |
| 9.a | 777 | 777 | 345 | $7102 / 2036$ |  |
| 9.a.N | 1206 | 1206 | 67 | $3613 / 2995$ | $753 / 624$ |
| 14.a | 174 | 174 | 0 | 0 | 0 |
| 14.b | 39263 | 39263 | 18 | $2489 / 63$ | $194 / 5$ |
|  |  |  |  |  |  |

In general, areas with insufficient sampling have relatively low levels of catch. The exception is Division 7.d from which 6 kt (mainly French) was caught which was not sampled. The number of age samples in southern fleets is disaggregated by area (included in Division 8.c total)

### 8.4 Catch Data

### 8.4.1 ICES Catch Estimates

The total ICES estimated catch for 2017 was 1155944 t , an increase of 61878 t on the estimated catch in 2016. Catches increased substantially from 2006-2010 and have averaged 1089 kt since from 2011.

The combined 2017 TAC, arising from agreements and autonomous quotas, amounts to 1194000 t ). The ICES catch estimate ( 1155944 t ) represents a slight undershoot of this. The combined fishable TAC for 2018, as best ascertained by the Working Group (see Section 8.1), amounts to 1000559 t .

Catches reported for 2017 and in previous Working Group reports are considered to be best estimates. In most cases, catch information comes from official logbook records. Other sources of information include catch processors. Some countries provide information on discards and slipped catch from observer programs, logbooks and compliance reports. In several countries discarding is illegal. Spanish data is based on the official data supplied by the Fisheries General Secretary (SGP) but supplemented by scientific estimates which are recorded as unallocated catch in the ICES estimates.

The text table below gives a brief overview of the basis for the ICES catch estimates.

| Country | OfFicial Log Book | Other Sources | DisCard Information |
| :--- | :--- | :--- | :--- |
| Denmark | Y (landings) | Y (sale slips) | Y |
| Faroe ${ }^{1}$ | Y (catches) | Y (coast guard) | NA |
| France | Y (landings) |  | Y |
| Germany | Y (landings) |  | Y |
| Greenland | Y (catches) | Y (sale slips) | Y |
| Iceland $^{1}$ | Y (landings) |  | NA |
| Ireland | Y (landings) |  | Y |
| Netherlands $^{\text {Norway }}{ }^{1}$ | Y (landings) | Y | Y |
| Portugal | Y (catches) |  | NA |
| Russia ${ }^{1}$ |  | Y (sale slips) | Y |
| Spain | Y (catches) |  | NA |
| Sweden | Y | Y | Y |
| UK (landings) |  | N |  |

${ }^{1}$ For these nations a discarding ban is in place such that official landings are considered to be equal to catches.

The Working Group considers that the estimates of catch are likely to be an underestimate for the following reasons:

- Estimates of discarding or slipping are either not available or incomplete for most countries. Anecdotal evidence suggests that discarding and slipping can occur for a number of reasons including high-grading (larger fish attract a premium price), lack of quota, storage or processing capacity and when mackerel is taken as by-catch.
- Confidential information suggests substantial under-reported landings for which numerical information is not available for most countries. Recent work has indicated considerable uncertainty in true catch figures (Simmonds et al., 2010) for the period studied.
- Estimates of the magnitude and precision of unaccounted mortality suggests that, on average for the period prior up to 2007, total catch related removals were equivalent to 1.7 to 3.6 times the reported catch (Simmonds et al., 2010).
- Reliance on logbook data from EU countries implies (even with $100 \%$ compliance) a precision of recorded landings of $89 \%$ from 2004 and $82 \%$ previous to this (Council Regulation (EC) Nos. 2807/83 \& 2287/2003). Given that over reporting of mackerel landings is unlikely for economic reasons; the WG considers that the reported landings may be an underestimate of up to $18 \%$ ( $11 \%$ from 2004), based on logbook figures. Where inspections were not carried out there is a possibility of a $56 \%$ under reporting, without there being an obvious illegal record in the logsheets. Without information on the percentage of the landings inspected it is not possible for the Working Group to evaluate the underestimate in its figures due to this technicality. EU landings represent about $65 \%$ of the total estimated NEA mackerel catch.
- The accuracy of logbooks from countries outside the EU has not been evaluated by WGWIDE. Monitoring of logbook records is the responsibility of the national control and enforcement agencies.

The total catch as estimated by ICES is shown in Table 8.4.1.1. It is broken down by ICES area group and illustrates the development of the fishery since 1969.

## Discard Estimates

With a few exceptions, estimates of discards have been provided to the Working Group for the ICES Subareas and Divisions 6, 7/8.a,b,d,e and 3/4 (see Table 8.4.1.1) since 1978. Historical discard estimates were revised during the data compilation exercise undertaken for the 2014 benchmark assessment (ICES, 2014). The Working Group considers the estimates for these areas are incomplete. In 2017, discard data for mackerel were provided by The Netherlands, France, Germany, Ireland, Spain, Portugal, Greenland, Denmark, England, Scotland and Sweden. Total discards amounted to 2832 t from these nations (mainly Spain and France). The German, Dutch, Irish and Portuguese pelagic discard monitoring programmes did not record any instances of discarding of mackerel. Estimates from the other countries supplying data include results from the sampling of demersal fleets.

Age-disaggregated data was limited but data available indicates that, in Divisions 8.a, $8 . \mathrm{b}$ and 8.c the majority of discarded fish were aged 0 to 3. In Division 9.a the majority of the discarded fish were 0 group.

Discarding of small mackerel has historically been a major problem in the mackerel fishery and was largely responsible for the introduction of the south-west mackerel box. In the years prior to 1994, there was evidence of large-scale discarding and slipping of small mackerel in the fisheries in Division 2.a and Sub-area 4, mainly because of the very high prices paid for larger mackerel $(>600 \mathrm{~g})$ for the Japanese market. This factor was put forward as a possible reason for the very low abundance of the 1991 year class in the 1993 catches. Anecdotal evidence from the fleet suggests that since 1994, discarding/slipping has been reduced in these areas.

In some of the horse mackerel directed fisheries, e.g. those in Subareas 6 and 7, mackerel is taken as by-catch. Reports from these fisheries have suggested that discarding may be significant because of the low mackerel quota relative to the high horse mackerel quota, particularly in those fisheries carried out by freezer trawlers in the fourth quarter. The level of discards is greatly influenced by the market price and by quotas.

### 8.4.2 Distribution of Catches

A significant change in the fishery took place between 2007 and 2009 with a greatly expanded northern fishery becoming established, and maintained to the present. Of the total catch in 2017, Norway accounted for the greatest proportion (19\%) followed by Scotland ( $16 \%$ ), Iceland ( $14 \%$ ), Russia ( $12 \%$ ) and Faroe ( $9 \%$ ). In the absence of an international agreement, Faroe, Greenland, Iceland and Russia declared unilateral quotas in 2017. Russia and Iceland both had catches over 100 kt with Faroes catching 99 kt. Greenlandic catches accounted for 46 kt of the total. Scotland had catch in excess of 100 kt and Ireland caught almost 86 kt. Germany, Netherlands, Spain, Denmark, France and England had catches of the order of $20-50 \mathrm{kt}$.

In 2017, catches in the northern areas (Subareas 2, 5, 14) amounted to 603869 t (see Table 8.4.2.1), an increase of 40366 t on the 2016 catch. Icelandic, Norwegian and Russian catches were all over 100 kt . Catches from Division 2.a accounted for $40 \%$ of the total catch in 2017. All the Russian catch in 2017 was taken in Division 2.a with Greenlandic catches taken further east into Division 2.a than in 2016. The wide geographical distribution of the fishery noted in previous years has continued.

The time series of catches by country from the North Sea, Skagerrak and Kattegat (Subarea 4, Division 3.a) is given in Table 8.4.2.2. Catches in 2017 amounted to 269804 t , an increase on 2016 ( 21193 t ). The majority of the catch is from Subarea 4 with small catches were also reported in Divisions 3.a-d.
Catches in the western area (Subareas 6, 7 and Divisions 8.a,b,d and e) increased slightly to 249229 t with most of the traditional fishing nations catching an increased proportion of their total catch in this area, likely due to the timing of the spawning migration. These catches are detailed in Table 8.4.2.3.

Table 8.4.2.4 details the catches in the southern areas (Divisions 8.c and 9.a) which are taken almost exclusively by Spain and Portugal. The reported catch of 33042 t represents a decrease from 2016. The catch is close to the long-term average.

The distribution of catches by quarter (\%) is described in the text table below:

| YEAR | Q1 | Q2 | Q3 | Q4 | YEAR | Q1 | Q2 | Q3 | Q4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 28 | 6 | 26 | 40 | 2004 | 37 | 6 | 28 | 29 |
| 1991 | 38 | 5 | 25 | 32 | 2005 | 46 | 6 | 25 | 23 |
| 1992 | 34 | 5 | 24 | 37 | 2006 | 41 | 5 | 18 | 36 |
| 1993 | 29 | 7 | 25 | 39 | 2007 | 34 | 5 | 21 | 40 |
| 1994 | 32 | 6 | 28 | 34 | 2008 | 34 | 4 | 35 | 27 |
| 1995 | 37 | 8 | 27 | 28 | 2009 | 38 | 11 | 31 | 20 |
| 1996 | 37 | 8 | 32 | 23 | 2010 | 26 | 5 | 54 | 15 |
| 1997 | 34 | 11 | 33 | 22 | 2011 | 22 | 7 | 54 | 17 |
| 1998 | 38 | 12 | 24 | 27 | 2012 | 22 | 6 | 48 | 24 |
| 1999 | 36 | 9 | 28 | 27 | 2013 | 19 | 5 | 52 | 24 |
| 2000 | 41 | 4 | 21 | 33 | 2014 | 20 | 4 | 46 | 30 |
| 2001 | 40 | 6 | 23 | 30 | 2015 | 20 | 5 | 44 | 31 |
| 2002 | 37 | 5 | 29 | 28 | 2016 | 23 | 4 | 44 | 29 |
| 2003 | 36 | 5 | 22 | 37 | 2017 | 24 | 3 | 45 | 28 |

The quarterly distribution of catch in 2017 is similar to recent years (since 2010) with the northern summer fishery in Q3 accounting for the greatest proportion of the total catch.

Catches per ICES statistical rectangle are shown in Figures 8.4.2.1 to 8.4.2.4. It should be noted that these figures are a combination of official catches and ICES estimates and may not indicate the true location of the catches or represent the location of the entire stock. These data are based on catches reported by all the major catching nations and represents almost the entire ICES estimated catch.

- First quarter 2017 (272 514 t - 24\%)

The distribution of catches in the first quarter is shown in Figure 8.4.2.1. The quarter 1 fishery is similar to that in previous years with the Scottish and Irish pelagic fleets targeting mackerel in Divisions 6.a, 7.b and 7.j. Substantial catches are also taken by the Dutch owned freezer trawler fleet. The largest catches were taken in Division 6.a, as in recent years. The Spanish fisheries also take significant catches along the north coast of Spain during the first quarter.

- Second quarter 2017 (39 972 t-3\%)

The distribution of catches in the second quarter is shown in Figure 8.4.2.2. The quarter 2 fishery is traditionally the smallest and this was also the case in 2017. The most significant catches where those in Division 8.c and at the start of the summer fishery in northern waters by Icelandic, Norwegian and Russian fleets.

- Third quarter 2017 ( 515346 t - $45 \%$ )

Figure 8.4.2.3 shows the distribution of the quarter 3 catches. Large catches were taken throughout Divisions 2.a (Russian, Norwegian vessels), 4.a (Norwegian, Scottish vessels), 5.a (Icelandic vessels). Catch was also taken in Division 14.b in quarter 3.

- Fourth quarter 2017 (328 112 t - 28\%)

The fourth quarter distribution of catches is shown in Figure 8.4.2.4. The summer fishery in northern waters has largely finished although there are substantial catches reported in the southern part of Division 2.a. The largest catches are taken by Norway, Scotland and Ireland around the Shetland Isles and along the north coast of Scotland. The pattern of catches is very similar to that reported in recent years.
ICES cannot split the reported mackerel catches into different stock components because there is no clear distinction between components upon which a split could be determined. Mackerel with a preference for spawning in the northeast area, including the North Sea, cannot presently be identified morphometrically or genetically (Jansen and Gislason, 2013). Separation based on time and area of the catch is not a precise way of splitting mackerel with different spawning preferences, because of the mixing and migration dynamics including inter-annual (and possibly seasonal) variation of the spawning location, combined with the post-spawning immigration of mackerel from the south-west where spawning ends earlier than in the North Sea.

### 8.4.3 Catch-at-Age

The 2017 catches in number-at-age by quarter and ICES area are given in Table 8.4.3.1. This catch in numbers relates to a total ICES estimated catch of 1155944 t . These figures have been appended to the catch-at-age assessment table (see Table 8.7.1.2).

Age distributions of commercial catch were provided by Denmark, England, Germany, Greenland, Faroes, Iceland, Ireland, the Netherlands, Norway, Portugal, Russia, Scotland and Spain. There remain gaps in the age sampling of catches, notably for French (length samples were provided), Swedish and Northern Irish fleets.

Catches for which there were no sampling data were converted into numbers-at-age using data from the most appropriate fleets. Accurate national fleet descriptions are required for the allocation of sample data to unsampled catches.

The percentage catch numbers-at-age by quarter and area are given in Table 8.4.3.2.
Over $80 \%$ of the catch in numbers consists of 3 to 8 -year olds with all year classes between 2010 and 2014 contributing over $10 \%$ to the total catch by number.

There is a small presence of juvenile (age 0) fish within the 2017 catch. As in previous years catches from Divisions 8.c and 9.a have contained a proportion of juveniles.

### 8.5 Biological Data

### 8.5.1 Length Composition of Catch

The mean lengths-at-age in the catch per quarter and area for 2017 are given in Table 8.5.1.1.

For the most common ages which are well sampled there is little difference to recent years. The length of juveniles is traditionally rather variable. Lengths recorded in 2017 for 0 and 1 group mackerel are lower than those in 2016. The rapid growth of 0 -group fish combined with variations in sampling (in recent years more juvenile fish have been sampled in northern waters whereas previously these fish were only caught in southern waters) will contribute to the observed variability in the observed size of 0-group fish. Growth is also affected by fish density as indicated by a recent study which demonstrated a link between growth of juveniles and adults ( $0-4$ years) and the abundance of juveniles and adults (Jansen and Burns, 2015). A similar result was obtained for mature 3- to 8-year-old mackerel where a study over 1988-2014 showed declining growth rate since the mid-2000s to 2014, which was negatively related to both mackerel stock size and the stock size of Norwegian spring spawning herring (Ólafsdóttir et al., 2015).

Length distributions of the 2017 catches were provided by England, Faroes, France, Iceland, Ireland, Germany, Greenland, the Netherlands, Portugal, Russia, Scotland and Spain. The length distributions were available from most of the fishing fleets and account for over $90 \%$ of the catches. These distributions are only intended to give an indication of the size of mackerel caught by the various fleets and are used as an aid in allocating sample information to unsampled catches. Length distributions by country and fleet for 2017 catches are given in Table 8.5.1.2.

### 8.5.2 Weights at Age in the Catch and Stock

The mean weights-at-age in the catch per quarter and area for 2017 are given in Table 8. 5.2.1. There is a trend towards lighter weights-at-age for the most age classes (except 0 to 2 years old) starting around 2005 is continuing until 2013 (Figure 8. 5.2.1). This decrease in the catch mean weights-at-age seems to have stopped since 2013 and values for the last five years do not show any particular trend for the older ages (age 6 and older) and are slightly increasing for younger ages (ages 1 to 5 ). These variations in weight-at-age are consistent with the changes noted in length in Section 8. 5.1.

The Working Group used weights-at-age in the stock calculated as the average of the weights-at-age in the three spawning components, weighted by the relative size of each component (as estimated by the 2016 egg survey for the southern and western components and the 2017 egg survey for the North Sea component). Mean weights-at-age for the western component are estimated from Dutch, Irish and German commercial catch data, the biological sampling data taken during the egg surveys and during the Norwegian tagging survey. Only samples corresponding to mature fish, coming from areas and periods corresponding to spawning, as defined at the 2014 benchmark assessment (ICES, 2014) and laid out in the Stock Annex, were used to compute the mean weights-at-age in the western spawning component. For the North Sea spawning component, mean weights-at-age were calculated from samples of the commercial catches collected from Divisions $4 . a$ and $4 . b$ in the second quarter of 2017 and the biological samples collected during the 2017 North Sea mackerel egg survey. Stock weights for the southern component, are based on samples from the Portuguese and Spanish catch taken in Divisions 8.c and 9.a in the $2^{\text {nd }}$ quarter of the year. The mean weights in the three component and in the stock in 2017 are shown in the text table below.

As for the catch weights, the decreasing trend observed since 2005 for fish of age 3 and older seems to have stopped in 2013 and values in the last four years do not show any specific trend (except for weights of ages 2 to 7 which have been increasing, Figure 8.5.2.2).

|  | North SEA <br> Component | WESTERN <br> Component | Southern Component | NEA MACKEREL <br> 2017 |
| :---: | :---: | :---: | :---: | :---: |
| Age |  |  |  | Weighted mean |
| 0 |  |  | 0.084 | 0.000 |
| 1 | 0.275 | 0.196 | 0.218 | 0.058 |
| 2 | 0.266 | 0.232 | 0.252 | 0.204 |
| 3 | 0.343 | 0.270 | 0.299 | 0.237 |
| 4 | 0.370 | 0.303 | 0.308 | 0.278 |
| 5 | 0.390 | 0.299 | 0.327 | 0.308 |
| 6 | 0.402 | 0.331 | 0.361 | 0.308 |
| 7 | 0.401 | 0.374 | 0.387 | 0.338 |
| 8 | 0.443 | 0.390 | 0.395 | 0.377 |
| 9 | 0.435 | 0.426 | 0.414 | 0.426 |
| 10 | 0.459 | 0.427 | 0.440 | 0.430 |
| 11 | 0.489 | 0.490 | 0.536 | 0.494 |
| $12+$ | $6.7 \%$ | $83.0 \%$ | $10.3 \%$ |  |
| Component <br> Weighting |  |  |  |  |
| Number of <br> fish <br> sampled | 399 | 458 | 1691 |  |

### 8.5.3 Natural Mortality and Maturity Ogive

Natural mortality is assumed to be 0.15 for all age groups and constant over time.
The maturity ogive for 2017 was calculated as the average of the ogives of the three spawning components weighted by the relative size of each component calculated as described above for the stock weights. The ogives for the North Sea and Southern components are fixed over time. For the Western component the ogive is updated every year, using maturity data from commercial catch samples collected during the first and second quarters (ICES, 2014 and Stock Annex). The 2017 maturity ogives for the three components and for the mackerel stock are shown in the text table below.

A trend towards later maturation (decreasing proportion mature at age 2) has been observed from the mid-2000s to 2011. A change in the opposite direction has been observed since then and the maturity ogive in 2017 is comparable with the one observed in the mid-2000s (Figure 8.5.3.1).

| Age | North Sea | Western Component | Southern <br> Component | NEA Mackerel |
| :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0.12 | 0.02 | 0.10 |
| 2 | 0.37 | 0.81 | 0.54 | 0.75 |
| 3 | 1 | 0.96 | 0.70 | 0.94 |
| 4 | 1 | 1 | 1 | 1 |
| 5 | 1 | 1 | 1 | 1 |
| 6 | 1 | 1 | 1 | 1 |
| 7 | 1 | 1 | 1 | 1 |
| 8 | 1 | 1 | 1 | 1 |
| 9 | 1 | 1 | 1 | 1 |
| 10 | 1 | 1 | 1 | 1 |
| 11 | 1 | 1 | 1 | 1 |
| $12+$ | 1 |  |  | $10.3 \%$ |
| Component |  |  |  |  |
| Weighting | $6.7 \%$ |  |  |  |

### 8.6 Fishery Independent Data

### 8.6.1 International Mackerel Egg Survey

### 8.6.1.1 Survey Planning for the 2019 Northeast Atlantic survey

The last mackerel egg survey was carried out in the NEA mackerel spawning areas in 2016 and a presentation with the final results were given during the WGWIDE meeting by the survey coordinator in 2017 (ICES, 2017b).

The ICES Working Group on Mackerel and Horse Mackerel Egg Surveys (WGMEGS) met in Dublin in April 2018 to plan the international mackerel and horse mackerel egg survey in 2019. The nations participating in the 2019 survey will be Portugal, Spain, UK Scotland, Ireland, The Netherlands, Germany, Norway, Iceland and the Faroe Islands.

The 2019 survey will be based on seven regular sampling periods. Additional information collated from summer surveys undertaken in 2017/2018 (Section 8.6.1.2) shows that mackerel spawning does only take place northwards the Faroe Islands if the temperature is higher than $8.5^{\circ} \mathrm{C}$ at 20 m depth. In addition, in 2018 summer survey was successful in delineating a zero-spawning boundary in the region encompassing Hatton Bank, the South Iceland Basin and all the way up to the Iceland Shelf.

The provisional survey plan of the 2019 mackerel and horse mackerel egg survey, as agreed during last the WGMEGS meeting (ICES, 2018c), is presented in Table 8.6.1.1.1.

In preparation for the 2019 survey a workshop dealing with egg identification and staging will take place during October 2018 in Bremerhaven. Procedures for fecundity and atresia estimation will be standardized and training conducted at the fecundity workshop to be held in IJmuiden in November 2018.

### 8.6.1.2 Results of the 2018 additional Mackerel Egg Survey in the northern survey area

The WGMEGS has been observing the offshore westwards and northwards expansion of the mackerel spawning area since 2007. In addition, results from the most recent
triennial MEGS survey in 2016 provided evidence that peak spawning of NEA mackerel had moved away from the traditional hotspots between the Bay of Biscay and the Porcupine Bank and instead was dispersed over a large swathe of open ocean, well away from the continental shelf to the West and Northwest of Scotland and importantly very close to the Northern and North-western survey boundary (Figure 8.6.1.2.1).

During the last mackerel benchmark in 2017, WGMEGS proposed several areas of additional work that required to be undertaken during the interim period $(2017,2018)$ and prior to the next triennial survey in 2019 (ICES, 2017a). The aim was to map the mackerel spawning activity within the North and North-western boundary areas and also hopefully delineate fully the mackerel spawning boundary, something that the triennial survey has hitherto been unable to deliver. The timing for these exploratory surveys/additional sampling was set at May/June.

The first exploratory egg survey was completed by Ireland during May/June of 2017. Results were presented at the last WGWIDE meeting (ICES, 2017b). The areas selected for survey were west of Hatton Bank, Southeast Iceland and the Faroes/Shetland channel. The results show that no stage 1 mackerel eggs were recorded in any of the sampled stations where the temperature at 20 m was less than 8 degrees Celsius (Figure 8.6.1.2.2). Therefore, the expected drop in temperature as the surveys proceeded northwards provides a physical barrier to mackerel spawning and the Northern boundary used by MEGS in 2016 should be relatively secure. However, potential mackerel spawning to the West of Hatton Bank and onto the South Icelandic Basin was less clear. This last area would be the target focus of the Scottish survey which was now scheduled for the same temporal period in 2018 (Burns et al, 2018).
During May/June of 2018 it was carried out the second exploratory survey on board a chartered Scottish fishing vessel (Altaire) with the objective of exploring the Northwestern boundary region and survey as far west as required until a zero spawning boundary was established. The survey deployed the Gulf 7 plankton sampler on a series of transects commencing on Rockall Bank and tracking East to West and vice versa heading steadily North up towards the Icelandic Shelf and also surveyed the West side of Iceland. In addition, there was support of the Nordic countries collecting extra plankton samples within this period during the International Ecosystem survey in the Norwegian Sea (IESNS) and Icelandic Spring Capelin surveys.

In this exploratory survey mackerel eggs were present in 49 of the 79 stations sampled with stage 1 mackerel eggs being identified in $60 \%$ of sampled stations. Virtually no mackerel eggs were recorded on stations where the temperature at 20 m was less than 8.5 degrees Celsius which is consistent with what is already known surrounding the temperatures tolerated by spawning mackerel. The survey successfully delineated the zero-spawning boundary in the Northwest (Figure 8.6.1.2.3). The relatively warmer temperatures observed on the flanks of Hatton Bank yielding moderate numbers of mackerel eggs whereas the colder water over the South Iceland Basin and also Northwards towards the Reykjanes Ridge being sufficiently cool as to provide the physical boundary and delivering few or zero mackerel eggs.

During 2018, additional plankton samples were collected by the Faeroe Islands, Iceland and also Norway during the IESNS survey. They covered a large swathe of ocean ranging from the East side of Iceland and North of Shetland to the Norwegian Coast. In addition, Iceland also collected 27 samples during their Capelin spring survey at the end of May and additional samples were also collected on the Icelandic Ecosystem surveys in the Nordic Seas in July-August (IESSNS) survey in mid-July. Analysis of the

IESNS samples concluded that none contained mackerel eggs (Figure 8.6.1.2.3). The same was found in both the Icelandic spring capelin survey samples and also those from the Icelandic IESSNS samples from July (Figure 8.6.1.2.4).

The survey results show that during May/June the spawning mackerel are avoiding crossing the cooler waters of the South Iceland Basin and instead are favouring the conditions on the Eastern side of the basin as they head North and certainly this is a widely held view. The total absence of mackerel eggs within the analysed IESNS samples is consistent with the results that were presented in 2017 and reaffirm the assessment that for the region stretching from the East coast of Iceland across to the Faroe/Shetland channel the existing Northern boundary surveyed by MEGS should be relatively secure with very little if any mackerel spawning taking place at that time of year at latitudes North of the Faroe Islands. No mackerel eggs were found in samples from any of the surveys where the recorded temperature at 20 m was less than 8 degrees Celsius. The significantly cooler sub-surface temperatures experienced in 2018 in the sampled areas around the Southern coast of Iceland had a significant impact on the abundance of mackerel eggs reported from the Icelandic Spring Capelin Survey samples with zero mackerel being reported in 2018. This was in a marked contrast to 2017 which recorded several stations with low to moderate densities of stage 1 mackerel eggs but with correspondingly warmer temperatures. It is entirely conceivable that this temperature anomaly may have had some impact regarding the distribution of spawning mackerel over the Hatton and South Iceland Basin region in 2018. However, the limited results reported from that area in 2017 provide some evidence that the pattern may not have been very different to that seen in 2018.

### 8.6.2 Demersal trawl surveys (Recruitment Index)

The index of survivors in the first autumn-winter (recruitment index) could not be updated due to input data quality issues in the ICES DATRAS system that had not been updated as recommended by WKWIDE 2017 (ICES, 2017a) and WGWIDE 2017 (ICES, 2017b). The outdated time series from WGWIDE 2016 (ICES, 2016a) was therefore used in the assessment. The assessment was therefore conducted without an index value for the 2016 and 2017 year classes and with the knowledge of an upcoming revision of the time series when the data quality issues has been sorted out.
The following text describes the methods used in 2016 and the data quality issues.

## The data and the model

An index of survivors in the first autumn-winter (recruitment index) was derived from a geostatistical model fitted to catch data from bottom trawl surveys conducted during autumn and winter. A complete description of the data and model can be found in Jansen et al. (2015) and the Stock Annex.

The data were compiled from several bottom trawl surveys conducted between October and March from 1998-2016 by research institutes in Denmark, England, France, Germany, Ireland, Netherlands, Norway, Scotland and Sweden. Surveys conducted on the European shelf in the first and fourth quarters are collectively known as the International Bottom Trawl Survey (IBTS). All surveys sample the fish community on the continental shelf and upper shelf slope. IBTS Q4 covers the shelf from the Bay of Biscay to North of Scotland, excluding the North Sea, while IBTS Q1 covers the shelf waters from north of Ireland, around Scotland, the North Sea, Skagerrak and Kattegat.

Trawl operations during the IBTS have largely been standardized through the relevant ICES working group (ICES, 2013a). Furthermore, the effects of variation in wing-
spread and trawl speed were included in the model (Jansen et al., 2015). Trawling speed was generally $3.5-4.0$ knots, and trawl gear is also standardized and collectively known as the Grande Ouverture Verticale (GOV) trawl. Some countries use modified trawl gear to suit the particular conditions in the respective survey areas, although this was not expected to change catchability significantly. However, in other cases, the trawl design deviated more significantly from the standard GOV type, namely the Spanish BAKA trawl, the French GOV trawl, and the Irish mini-GOV trawl. The BAKA trawl had a vertical opening of only $2.1-2.2 \mathrm{~m}$ and was towed at only 3 knots. This was considered substantially less suitable for catching juvenile mackerel and, therefore, was excluded from the analysis. The French GOV trawl was rigged without a kite and typically had a reduced vertical opening, which may have reduced the catchability of pelagic species like mackerel. Catchability was assumed to equal the catchability of the standard GOV trawl because testing has shown that the recruitment index was not very sensitive to this assumption (Jansen et al., 2015). Finally, the Irish mini-GOV trawl, used during 1998-2002, was a GOV trawl in reduced dimensions which was accounted for by inclusion of the wing-spread parameter in the model.

A geostatistical log-Gaussian Cox process model (LGC) with spatiotemporal correlations was used to estimate the catch rates of mackerel recruits through space and time. The modelled average recruitment index (squared CPUE) surface was mapped in Figure 8.6.2.1. The time-series of spatially integrated recruitment index values was used in the assessment as a relative abundance index of mackerel at age 0 (recruits) - see Figure 8.6.2.2.

## Survey Coverage

The combined demersal surveys have insufficient spatial coverage in some areas that can be important for the estimation of age-0 mackerel abundance, namely: (i) Since 2011, the English survey (covering the Irish sea and the central-eastern part of the Celtic sea including the area around Cornwall) has been discontinued; (ii) the Scottish survey has not consistently covered the area around Donegal Bay; and (iii) the IBTS has observed high catch rates in some years at the north-eastern edge of the survey area (towards the Norwegian trench) in winter. It is therefore possible that some recruits are also overwintering on the other side of the trench along the south western shelf edge of Norway. Consequently, the Norwegian Sea IBTS (NS-IBTS) in first quarter (Q1) should be extended to include the south-western Norwegian shelf and shelf edge in proximity to the Norwegian trench.

## Data Quality

Errors in the input dataset have been detected since WGWIDE 2016. Data revisions by Scotland and Ireland were done before WGWIDE 2017, but for WGWIDE 2018 the ICES DATRAS system was not updated to deliver data and quality assurance reports as recommended by WGWIDE 2017 and WKWIDE 2017. It was therefore not possible to update the time series during the meeting. It is expected that the ICES datacentre will complete this work during autumn 2018, well before the next assessment (or intermediate benchmark), because significant progress was seen in the weeks before the meeting and during the meeting. The recommendations to ICES datacentre will therefore not be repeated this year.

This should facilitate a revision of the recruitment index in time for the 2019 assessment. For the update assessment WGWIDE 2018 used the time series from WGWIDE 2016 (Figure 8.6.2.2).

Mackerel samples collected on the EVHOE fourth quarter (Q4) survey are not aged. The current practice of applying age-length keys from Ireland and Scotland to catches in the more southern EVHOE survey is not ideal, because the mackerel growth during the first year is related to latitude (Jansen et al., 2013). WGWIDE therefore recommends that Ifremer (France) initiate aging of mackerel starting from Q4 2018.
Finally, WGWIDE encourage studies of vertical distribution and catchability of age-0 mackerel in the Q4 and Q1 surveys.

### 8.6.3 Ecosystem surveys in the Nordic Seas in July-August (IESSNS)

The IESSNS was successfully conducted in the summer of 2018 (Figure 8.6.3.1). Five vessels sampled 290 predetermined surface trawl stations in the period from June 30 to August 6 which covered an area of 2.8 mill. $\mathrm{km}^{2}$ which is the same as in 2017 (ICES, 2018a). At each surface trawl station, a standardized trawl (Multpelt832) is employed for $30-\mathrm{min}$ according to a standardize operation protocol which is designed to catch mackerel. Additionally, abundance of herring and blue whiting is measured using acoustic methods and backscatter is verified by trawling on registrations as needed. The aim is to establish an index for blue whiting and herring abundance to be used in stock assessment in a few years. The cruise report is available as a working document to the current report (ICES, 2018a) and a detailed survey description is in the Stock Annex.

IESSNS provides annual age-segregated index for mackerel abundance of which age classes 3-11 are used to tune the mackerel stock assessment (Table 8.6.3.1; Ólafsdóttir et al., 2017; ICES, 2017a).

Excluding the North Sea, the total swept area abundance index of mackerel in 2018 was estimated 16.9 billion individuals which is a decrease of $30 \%$ compared to 2017. Mackerel biomass index declined $40 \%$ between years (Figure 8.6.3.2). The discrepancy in decline of abundance index and biomass index is due to record high numbers of age- 1 and age- 2 mackerel and lower weight-at-age for these age classes in 2018 compared to 2017. The most abundant year classes were 2010, 2011, 2014, 2016 and 2017 respectively presenting $11 \%, 14 \%, 14 \%, 15 \%$, and $13 \%$ of the stock in numbers (Figure 8.6.3.3). The incoming 2017-year class has the largest age-1 index value recorded in IESSNS and is $150 \%$ larger than the incoming age- 1 cohort in 2017. Mackerel cohort internal consistency has improved by adding the 2018 survey data to the time series. Mackerel cohort internal consistency remained relatively high. Internal consistency is strong for ages 1 to 5 years ( $r>0.8$ ) and a fair/good internal consistency for ages 5 to 11 years ( $r>$ 0.5 ), except for 7-8 year old mackerel (Figure 8.6.3.4)

The North Sea (southward of latitude $60^{\circ} \mathrm{N}$ ) was included in the IESSNS for the first time in July 2018 and 39 predetermined surface trawl stations were sampled. The survey area was 0.25 mill. $\mathrm{km}^{2}$, and the estimate index for mackerel abundance was 2.2 billion individuals and the biomass index was 0.4 million tonnes. The North Sea survey areas is excluded for the calculations of the mackerel abundance index used in the assessment according to the 2017 benchmark (Ólafsdóttir et al., 2017; ICES, 2017a), hence the results are presented separately from the traditional survey north of latitude $60^{\circ} \mathrm{N}$.

### 8.6.4 Tag Recapture data

The Institute of Marine Research in Bergen has annually conducted tagging experiments on mackerel since 1968, both in the North Sea and to the west of Ireland during the spawning season May-June. However, only the information from mackerel tagged west of Ireland is used in the mackerel assessment, and only information on recaptures
of mackerel tagged with steel-tags until 2006. A new RFID tagging method from 2011 onwards was accepted and used in the assessment based on the conclusions from the 2017 WKWIDE benchmark workshop (ICES, 2017a).

## Steel-tags

These tags have been recovered at metal detector/deflector gate systems installed at plants processing mackerel for human consumption. This system demanded a lot of manual work, paying for external personnel to stay at the plants during processing. Among the typical 50 fish deflected, the hired personal must find the tagged fish with a hand-hold detector and send the fish to IMR for analysis. This has been time consuming and expensive. Besides being used in present mackerel assessment model, the tagging data have also been used in estimates of mortality, and recently in estimation of spawning stock biomass, and further has the tagging data been valuable for understanding the migration of the mackerel (Tenningen et al., 2011).

## RFID tags

## General description of data

The radio-frequency identification (RFID) tagging project on NEA mackerel was initiated in 2011 at the Institute of Marine Research, Bergen (IMR) in Norway. RFID is a technology that uses radio waves to transfer data from an electronic tag, called an RFID tag, through a reader for the purpose of identifying and tracking the object. The new RFID tagging project has moved away from manual and expensive to an automatic and cost-effective scanning system.

During the period 2011-2016 as many as 353541 mackerel has been tagged with the new tags and 3337 of these tags have been recaptured (Table 8.6.4.1). This includes an experiment off the Norwegian Coast on young mackerel in September 2011 as well as three experiments carried out in August in Iceland 2015-2017, none of which is included as input data in the assessment. Data from the releases at the spawning grounds in May-June of Ireland and the Hebrides are the only data included in the assessment.

The RFID-tagged mackerel recaptured up to $1^{\text {st }}$ September 2018, came from 22 European factories processing mackerel for human consumption (Table 8.6.4.2). The project started with RFID antenna reader systems connected to conveyor belt systems at 8 Norwegian factories in 2012. Now there are 6 operational systems at 5 factories in UK (Denholm has 2 RFID systems) and 2 in Iceland. Norway has installed RFID systems at 8 more factories in 2017-2018, most of which with the purpose of scanning Norwegian spring spawning herring catches (IMR started tagging herring in 2016), but some also processing mackerel. More systems are also bought by Ireland (3), which up to now has been non-operational. Note also that in the current assessment data from the factories Sæby (Denmark), Lunar Freezing Frazerburgh (Scotland), Höfn (Iceland), Austevoll and Egersund (after 2013) in Norway are all excluded due to problems with efficiencies and low recapture rates. The factories having operational systems are all online on internet and RFID tagged mackerel recaptured by the systems are automatically updated in the central database in Bergen with date, time, and factory of location.

There is a web-based software solution and database that is used to track the different systems, import data on catch information, and biological sampling data of released fish and screened catches. Based on this information the system can estimate numbers released every year, and the concurrent numbers screened and recaptured over the next years (by year class), which is what is used in the assessment. The development of the tagging data time series is dependent on the work from each country's research
institutes, fisheries authorities or the industry its selves to provide additional data about catches screened through the RFID systems, such as total catch weight, position of catch (ICES rectangle), mean weight in catch, etc. Regular biological sampling of the catches landed at these factories is also needed. Altogether, these data are essential for the estimation of numbers screened per year class. Responsible scientists in Norway, Iceland, Faroes and Scotland has been following up the factories, and delivering the catch data and biological data. In the future, it is planned that annual workshops should occur prior to the assessment, where more scientists go through the new data being updated from new tagging experiments, as well as recaptures from all previous experiments, and undertake analyses of the trends in the data outside of the assessment model, see suggestions to terms of reference for such an annual workshop at the end of this section.

## Trends and bias concerns in the RFID tag-recapture data

The way the tagging data is used in the SAM assessment model is more of a raw data format, rather than an abundance index adding one new number per age per year (one line in a table). What is used is number released every year of a year class, and the numbers scanned and recaptured every year of the same year classes in all the years after release. The model is estimating the size of a year class in the release year, based on data from all recapture years. This means for example that the recaptures from the 2011 experiment in year 2017, in fact influences the prediction of the abundance in 2011, meaning that the prediction of 2011 abundance may change over time with more recapture years. This is very different from other typical indexes of abundance normally used in assessments.

The way the tagging data are handled also means that there is no index presented really showing the trends in the data, such as with the egg survey and the IESSNS trawl survey. However, this is possible by estimating the abundance/biomass in the release year using the Petersons model ( $\mathrm{N}=$ numbers released/numbers recaptured*numbers scanned). During WGWIDE 2018 several results were shown to demonstrate the trends in the RFID tag data based on Peterson's estimation, some of which indicated biases in the data that could influence the assessment. In the following the main results will be described. All estimates are scaled to the $10 \%$ survival also used as scaling in SAM, not taking into account the mortality happening over the year (which also currently is not being taken into account in SAM).

When only estimating the biomass in release year based on recaptures the first year after release, one is able to follow the trend from 2011-2017 (Figure 8.6.4.1), where the estimate in 2017 is based only on quarter 1 recaptures in 2018. The trends in biomass of age2+, 3+ and 4+ mackerel show some similarities, all decreasing in the end towards 2017. However, the age2+ index seems to have a jump in 2013, suggesting some noise in the data when the large 2010-2011 year classes are entering in the tagging experiments at the ages 2-3 years.

However, given that SAM takes into account all recapture years, it is important to look at potential changes in the estimates related to recaptures at longer times after release than 1 year. The results when estimating the trends in biomass of age $2^{+}$, but based on different numbers of years out (1-6 years), clearly show a trend for release years 20112012 that is indicating a bias in the data (Figure 8.6.4.2), the estimates increasing heavily with the numbers of years out. This is not according to the assumption in using tag data for abundance estimation, where it is expected that it should be stable when the fish has mixed in the stock. The bias is not so clear for the years after 2012.

When looking closer into the bias in estimating biomass by age groups from tagging data with numbers of years out, from all release years (Figure 8.6.4.3), it indicates that the problem is highest in the young fish, where the change over time is highest. Especially, this can be seen for the strong 2010-2011 year classes entering the tagging data.

There seems to be some change in the estimates from tagging data happening after 2012, and it is important to notice that this also corresponds to a large change in the distribution and abundance of catches scanned for tags (Figure 8.6.4.4). From 2014 onwards Icelandic, Faroes and Scottish factories really contributed to a tripling of the scanned biomass, and a change with a broader distribution of scanned catches in the Norwegian Sea and eastwards to Iceland during quarter 3-4, as well as a significant increase in quarter 1 along the British Isles and Ireland. This change alone could have caused changes in the ways the tagging data effect the assessment, especially if the recapture rates in different areas/seasons vary, according to lack of mixing of tagged fish, or according to mortality happening between seasons (for instance between quarter 1 and quarter 4 catches), which is not taken account for in SAM assessment today.

To check for potential area/season effects on the tag recapture data, the data were reanalysed, based on a splitting in 4 different areas/seasons (Figure 8.6.4.5). The results when looking at trends in the biomass of age 2+ mackerel, when estimated based on recaptures from the 4 areas in the year 1 after release, shows more noise and variation, but still the same trend towards lower biomass in 2016-2017 (Figure 8.6.4.6). The area that seems to stick out is the central Norwegian Sea, which tend to have higher estimates than the others, indicating lower recapture rates, which could suggest a problem with mixing. When looking more detailed into this potential problem, estimating the biomass by age for each release year, based on each of the 4 recapture areas, and different numbers of years after release (Figure 8.6.4.7), even more of the variability in the data are shown. One thing to notice is the noise in the data in 2013, especially coming from the estimate of the 2010 year class based on recaptures from area in the central Norwegian Sea.

## Exploratory runs in SAM related to concerns of bias issues in the RFID tagging data

Based on the results above it was decided to do some exploratory runs in SAM to look for sensitivity to the inclusion of different ranges of age groups, different numbers of years of recaptures included after releases, and different areas/seasons.

- Using ages $2+, 3+, 4+$
- Using years out=all years or years out <3
- Using the new tag data set split into 4 areas/seasons

The results of these exploratory runs are shown under SAM assessment results (see Section 8.7.4).

Alternative use of tag data in the assessment - use an index?
In WGWIDE 2018 there was a lot of discussion with regard to the handling of the tag data in SAM. One point raised was that the "raw data" format used for tag data, results in a lot more data, increasing for every year, and how this is handled for instance with regard degrees of freedom. It was discussed in the group that a simpler use of the tag data, in terms of a regular abundance index, perhaps would be more appropriate. At least this would open up for an easier way for other assessment models to use the data for comparisons, especially given the circumstances of the current assessment, where the tag data seems to get a very high weight. There are several ways to make such and index, one attempt is shown in Table 8.6.4.3 and Figure 8.6.4.8. Here it is assumed that
by only including recapture data from the two first years after release (YearsOut=1-2), the estimation in all release years are treated in the same way, and a potential bias with reduced recapture rates with increasing numbers of years after a release is reduced. It is also assumed that data from fish at ages 2-3 are more uncertain, noisier, for instance the 2010-2011 year classes tagged at ages 2-3 years seems noisy. Hence, only data from ages 4-12 were included. For the sake of comparison, the data were also scaled down to the $10 \%$ survival used in SAM assessment. This index is something that can be tried out in SAM and other models as an alternative way to use the data, at least for exploration.

## Regarding the issue with low survival rate in the RFID tag data

Work is being done to try understanding the different estimated scaling parameters on the 'old' steel tag (survival= $40 \%$ ) vs the new RFID tag times series (survival=10\%), that cannot be explained by suggested bias issues in the new RFID-time series, but actual change in tagging mortality, tag loss or detection-efficiencies at factories. This needs focus and attention as it is not understood by the responsible taggers who evaluate every single fish prior to tagging, nor the responsible scientists.

Some work is already done, such as testing off detection efficiencies at the factories. However, there is clearly need for more testing, several times over the season at all factories. This is something that needs priority, and the plan is to carry out extensive testing until next WGWIDE meeting, or potentially prior to an intermediate benchmark at an earlier date. We need to make sure if the efficiency is stable at high levels, or to adjust for potential variability if this should be the result of extensive testing.

With regard to testing of tagging mortality, some tests are also carried out already. One test is that Iceland in fact has started their own experiments, where the handling of the fish is a little different than in the experiments of Ireland, and where the fish itself perhaps is less sensitive as it is not in a spawning condition as it is off Ireland. However, a comparison in biomass estimates by age and totally between the two experiments in 2016 based on recaptures in 2017 (Figure 8.6.4.9), showed overlapping estimates. This suggests equal survival rates from the two experiments despite the different handling, and condition of the fish. There has also been experiments of Iceland in 2017 and 2018, and it will be of value to follow and compare with the experiments off Ireland in the years to come, to follow up on the discussion of low survival rate on RFID tagged fish estimated by SAM.
Another test for evaluating if the change of handling of the mackerel from the old steel tagging to the new RFID tagging, is that in 2017 a proportion of the tagged fish was handled in the exact same way as used for the steel tags; meaning that: (i) using manual jigging instead of automatic jigging machines; and (ii) using old rectangular tanks for keeping the fish compared with circular tanks, and releasing the fish directly to the sea on starboard side instead of through pipes on the port side. The difference is only in the tag type used, and to some extent the placement of tags; meaning that the old steel tags were inserted into abdomen of the fish, if not in a spawning stage, and into the muscle of the fish when in a spawning stage, as compared with RFID tags, which always are inserted into the abdomen. The decision to always insert the RFID tags into abdomen is to avoid that tags are going all the way to the consumer. The result from the 2017 experiments may help understanding if the handling of the fish is a reason to the differences in survival rates estimated by SAM, but some time with recaptures is necessary prior to conclusion from this experiment.

Another alternative is to carry out large scale tagging experiments at sea, releasing tagged fish into large sea pens, floating around for a period, after which the mortality
could be assessed. Such experiments are possible to conduct, but they will not necessarily show realistic mortalities, as swimming in a pen is not comparable to swimming in the open ocean. Still, such experiments may increase the understanding of the low estimated survival rate, and it is clearly something that should be considered in the future. It must, however, be emphasized that all previous experiments on tagging mortality on mackerel are not realistic with regard to the actual mortalities that are happening out at sea. To underline this, IMR has carried out experiments on both herring and mackerel, with close to zero mortality due to tagging process, when the fish were in really good condition under low stress prior to tagging. The conclusion is that most of the mortality happening is caused by all the handling and stress caused from being hooked with jigging, until the release at sea, not the tag insertion itself. This means that realistic experiments must be carried out under the same conditions normally experienced when tagging in the open ocean with the vessels currently used.

Suggested terms of references for an annual ICES workshop on tagging data
As mentioned above, there is need for an annual workshop dealing with the tagging data for mackerel, but also for Norwegian Spring Spawning herring where tagging started in 2016. Below are the suggested terms of reference for such a working group that should preferably meet in spring prior to the WGWIDE assessment.

- Update the tagging database with all new data needed (catch data and biological data) and carry out estimations needed for updating the tag data table used in the SAM assessment.
- Quality assurance of the tag data table, hereunder to consider if adjustments are needed in tag data table, such as removal of data previously used from factories with low efficiency or alternative use of biological data (such as ALKs) to estimate numbers released and scanned by age.
- Carry out analyses of the trends (indexes of abundance by age and biomass) in the tag data outside the SAM model that can be presented to WGWIDE.
- Plan experiments and carry out analyses that may be used to shed light on the low survival rate estimated for the RFID tags, such as proper testing and control of detection efficiency at factories, survival experiments, special tagging experiments.
- Prepare a full report of the results from the workshop to be presented at WGWIDE.


### 8.6.5 Other surveys

### 8.6.5.1 International Ecosystem survey in the Norwegian Sea (IESNS)

After the mid-2000s an increasing amount of mackerel has been observed in catches in the Norwegian Sea during the combined survey in May (IESNS) targeting herring and blue whiting (Rybakov et al., 2016; 2017). The spatial distribution pattern was slightly reduced in 2018, where mackerel was caught within a more limited area and in fewer trawl stations of the Norwegian Sea compared to 2017 (Rybakov et al., 2017; ICES, 2018b). Mackerel at age 2 (mean length 26.4 cm ) was most numerous in the combined samples and amounted to $26 \%$, followed by age 1 (17\%) and age 5 (13\%) (ICES 2018b).

The mackerel distribution was further east in 2018 compared to in 2017. In 2018, the northernmost mackerel catch was at $70^{\circ} \mathrm{N}$ and the westernmost catch was at $2^{\circ} \mathrm{W}$. In 2017, the northernmost mackerel catch was at $71^{\circ} \mathrm{N}$ and the westernmost catch was at $10^{\circ} \mathrm{W}$. There was a less pronounced distribution of 1-year old mackerel found in 2018
compared to in 2017. There was still a northerly distribution of 1-year old mackerel in the northeast, whereas it was indicated that the 2017-year class also was the most dominant one year later, now as 2-year old mackerel in 2018. The IESNS survey provide valuable although limited quantitative information can be drawn. This acoustic based survey is not designed to monitor mackerel, and do not provide proper mackerel sampling in the vertical dimension, and also involve too low trawl speed for representative sampling of all size groups of mackerel. The trawl hauls are mainly targeting acoustical registrations of herring and blue whiting during the survey in May (IESNS).

### 8.6.5.2 Acoustic estimates of mackerel in the Iberian Peninsula and Bay of Biscay (PELACUS)

Due to the participation in the International Blue Whiting Spawning Stock Survey (IBWSS), PELACUS 0318, was started a little bit later than previous years (25/03 instead $16 / 03$ ), and the area was clockwise steamed, from the inner part of the Bay of Biscay to the Spanish -Portuguese border, thus contrary to the normal procedure (Carrera et al., 2018a,b). Weather conditions were adverse, with a continuous low-pressure fronts with dominant SW/W winds and swell of about 4 m height, resulted in an important haline front all around the surveyed area due to the river run-offs, and a poleward current with clear influence up to $6^{\mathrm{a}} 30^{\prime} \mathrm{W}$ (Galician waters). These conditions might have been an important influence in both aggregation pattern and spatial distribution in most of the fish species. In the case of mackerel, the distribution area was mainly restricted to coastal waters ( $<150 \mathrm{~m}$ depth) and mainly occurring in thick bottom layers. Together with mackerel, other swim bladder species were also found in these layers, as revealed by the frequency response done in those echotraces. The increase towards higher frequencies was lower than expected. Ground truth fishing stations confirmed this presence, although mackerel accounted up to $95 \%$ of the total catch in number. For this reason, instead of direct allocation, the Nakken and Dommasnes (1975) method for multiple species was used to split backscattering energy into those fish species caught at the ground truthing trawl hauls.

The bulk of the distribution, as in previous years, was located just in the middle of the Cantabrian Sea (Cape Peñas), extending throughout the surveyed area (Figure 8.6.5.2.1). A total of 557 thousand tonnes, corresponding to 1640 million fish were estimated, most of them, as expected, in central Cantabrian Sea (Figure 8.6.5.2.2, Tables 8.6.5.2.1-2). This is similar to that assessed in 2017 ( 548 thousand tonnes corresponding to 1777 million fish). As observed in previous years, only few individuals younger than 5 years were estimated (less than $10 \%$ in weight, $14 \%$ in number) Age group 6 was dominant ( $25 \%$ ). Mean length was 36.1 cm with a mean weight of 318.3 g , without any significant change in mean length nor in length distribution along the surveyed area.
On the other hand, this year mackerel egg collected by CUFES were counted and staged. 98\% (364 of 373 station- each of them corresponding to 3 nmi on average-) resulted positive for mackerel eggs, with a mean of 248 egg per station ( $24 \mathrm{eggs} / \mathrm{m}^{3}$ ). These figures are much higher than those collected at the Porcupine Sea Bight, where only few eggs were counted, with only $0.62 \mathrm{eggs} /$ station ( $0.05 \mathrm{eggs} / \mathrm{m}^{3}$ ) (Figure 8.6.5.2.3).

### 8.7 Stock Assessment

### 8.7.1 Update assessment in 2017

NEA mackerel was classed as an update assessment this year. The update assessment was carried out by fitting the state-space assessment model SAM (Nielsen and Berg,
2014) using the $R$ library stock assessment (downloadable at install_github("fishfollower/SAM/stockassessment", ref="mack")) and adopting the configuration described in the Stock Annex.

The assessment model is fitted to catch-at-age data for ages 0 to 12 (plus group) for the period 1980 to 2017 (with a strong down-weighting of the catches for the period 19801999) and three surveys: (i) the SSB estimates from the triennial Mackerel Egg survey (every three years in the period 1992-2016); (ii) the recruitment index from the western Europe bottom trawl IBTS Q1 and Q4 surveys (1998-2015, not updated for the last 2 years); and (iii) the abundance estimates for ages 3 to 11 from the IESSNS survey (2010, 2012-2018). The model also incorporates tagging-recapture data from the Norwegian tagging program (for fish recaptured between 1980 and 2005 for the steel tags time series, and fish recaptured between 2012 and 2017 for the radio frequency tags time series).

Fishing mortality-at-age and recruitment are modelled as random walks, and there is a process error term on ages 1-11.

The differences in the new data used in this assessment compared to the last year's assessment were:

- No update of the IBTS recruitment index was available and the time series did not include any 2016 and 2017 estimates (see Section 8.6.2).
- Addition of the 2018 survey data in the IESSNS indices.
- Addition of the 2017 catch-at-age, weights-at-age in the catch and in the stock and maturity ogive, proportions of natural and fishing mortality occurring before spawning.
- The inclusion of the tag recaptures from 2017, and minor revision in the tagging recapture data set for the RFID tagging program for the earlier recapture years (differences less than $1 \%$ in the recapture rates).

Input parameters and configurations are summarized in Table 8.7.1.1. The input data are given in Tables 8.7.1.2 to 8.7.1.9. Given the size of the data base the tagging data are not presented in this report, but are available on www.stockassessment.org in the data section (files named tag.dat and tag3.dat).

### 8.7.2 Model diagnostics

## Parameter estimates

The estimated parameters and their uncertainty estimates are shown in Table 8.7.2.1 and Figure 8.7.2.1. The model gives a good fit to the catch data (lowest observation standard deviation). The observation standard deviations for the egg survey is also low, indicating a good fit to this survey. The observations standard deviations for the recruitment index and the IESSNS surveys ages 4 to 11 are higher indicating that the assessment gives a lower weight to the information coming from these surveys. The IESSNS age 3 has a very low weight in the assessment (high observation standard deviation). Overdispersion of the tag recaptures is not directly comparable with observation standard deviation, but has the same meaning. The model assigns a similar overdispersion to the steel tag data and the RFID tag data.

The catchability of the egg survey is 1.37 , significantly larger than 1 , which implies that the assessment considers the egg survey index to be an overestimate. The catchabilities at age for the IESSNS increase from close to 1 for age 3 to 2.69 for age 7 and decreases slightly for older ages. Since the IESSNS index is expressed as fish abundance, this also
means that the assessment considers the IESSNS to provide over-estimated abundance values for the oldest ages. The post tagging survival estimate is higher for the steel tags (around $40 \%$ ) than for the RFID tags (around $10 \%$ ).

The process error standard deviation (ages 1-11) is moderate (lower than in previous assessments) as well as the standard deviation of the F random walk.

The catchability parameters appear to be estimated more precisely than the observation standard deviations, except for the catchability of the IESSNS at age 3 which has a higher standard deviation. Uncertainty on the tags post release survival is low. Uncertainty on the observation standard deviations is larger for the egg survey and the IESSNS age 3 than for the other survey indices. Uncertainty on the overdispersion of the RFID tag data is high.
The estimated AR1 error correlation structure for the observations from the IESSNS survey age 3 to 11 a high correlation between the errors of adjacent ages ( $r=0.82$ ), then decreasing exponentially with age difference (Figure 8.7.2.2.). This high error correlation implies that the weight of this survey in the assessment in lower than for a model without correlation structure, which is also reflects in the high observation standard deviation for this survey.

There are some strong correlations between parameter estimates (Figure 8.7.2.3):

- Catchabilities are positively correlated (especially for the IESSNS age 4 to 11), and negatively correlated to the survival rate for the RFID tags. This implies that the model cannot distinguish well between low catchabilities / high tag survival (larger stock) and high catchability / low survival (smaller stock).
- The observation variance for the IESSNS age 4-11 is positively correlated to the autocorrelation in the errors for these observations. This implies that when the model estimates highly correlated errors between age-groups, the survey is considered more noisy.
- The observation variance of the catches is negatively correlated to the variance of the fishing mortality random walk. This implies that when the model tends to consider the catches as more precise, this implies a more variable fishing mortality.

These correlations mean that the model is not able to estimate these parameters independently and may indicate that it is overparameterised.

## Residuals

The "one step ahead" (uncorrelated) residuals for the catches did not show any temporal pattern (Figure 8.7.2.4) except for 2014 for which they were mainly positive for 2014 (modelled catches lower than the observed ones). This may result from the random walk that constraints the variations of the fishing mortality, which prevents the model from increasing the fishing mortality suddenly (which probably happened given the sharp increase in the catches in 2014). Residuals for ages 0 and 1 are larger than for subsequent ages 2 to 10 . Residuals for ages 11 to 12 are also larger than for ages 2 to 10 . This suggest that decoupling the observation variance of the catches (for example by grouping age 0 and 1 , ages 2 to 10 and ages 11 and older) could be more appropriate. This has been investigated during the last benchmark assessment, but the model with decoupled observation variances gave a very tight fit to the recruitment index (observation standard deviation close to 0.05 ) and a very large observation standard deviation for the catches of ages 0 and 1 . WKWIDE 2017 regarded the tight fit to
the recruitment index as unrealistic and chose to retain the current model structure because there was insufficient time to continue with this analysis (ICES, 2017a). WGWIDE recommends that this work is prioritized during the next benchmark, because the problem with juvenile catches remained unsolved.

The residuals for the egg survey show a slight temporal pattern with negative residuals in the period 2001-2004 and followed by positive residuals for the period 2007-2013. The residual for the 2016 point is large and negative, indicating that the model has difficulties fitting to this low estimate, despite the overall decrease in the estimated SSB over the recent years.

Residuals for the IESSNS indices do not show any marked pattern, except the predominance of positive residuals for the two recent years (2016 and 2017) which indicate that the model does not agree with the high value of the survey observed for these 2 years. Residuals for the latest year are more balanced.

Residuals to the recruitment index show no particular pattern.
Finally, inspection of the residuals for the tag recaptures (Figure 8.7.2.5) did not show any sign of model misspecification. The only minor concern was for fish released at age 2 for which the predominance of positive residuals suggested that the post-release mortality for those fish may have been lower than for other ages (more tags return than expected). This issue is studied in more details in Section 8.7.4.

## Leave one out runs

In order to visualise the respective impact of the different surveys on the estimated stock trajectories, the assessment was run leaving out successively each of the data sources (Figure 8.7.2.6). All leave out one runs showed parallel trajectories in SSB and Fbar, except the run removing the tagging data which shows a different dynamics in the early period of the assessment (before 2000) and in the recent years (since the start of the RFID time series). Further inspection of the output of the run without the tagging data showed that the model is not able to estimate accurately the parameters (it resulted this year in a variance for the F random walk close to 0 , corresponding to constant F for the whole time series). This is explained by the fact that, without tagging data, the model has no information on the period prior to 2000, expect 3 egg survey points. The leave one out run excluding the tagging data should therefore be disregarded.

Removing the recruitment index had only on minor effect on the estimated stock trajectory. Removing the IESSNS resulted in lower SSB estimates and higher Fbar estimates for the period covered by the survey. On the opposite, removing the egg survey results in a larger estimated stock, exploited with a lower fishing mortality. These 2 surveys have a notable contribution to the assessment (even if the leave one out runs fall within the confidence intervals of the assessment using all data), and in a way, the final assessment seems to make a trade-off between the information coming from the IESSNS which lead to a more optimistic perception of the stock, and the information from the egg survey which suggest a more pessimistic perception of the stock. This conflict between the 2 surveys seem to have decreased compared to previous years, as the difference between the 2 leave one out runs is less pronounced this year than in the past.

The sensitivity of the assessment was tested for the RFID data alone in a separate analysis (Figure 8.7.2.7). Removing this source of data result in a very different perception of the development of the stock after 2012: the SSB in the assessment without RFID tag
data continues to increase to reach close to 5.5 million tonnes in 2015 before declining to 4 million tonnes in 2017, while the SSB in the assessment using the RFID tags decreases continuously since 2011 to reach levels close to 3 million tonnes in 2017. The influence of a single year of data for the RFID tags was also tested by removing the recaptures from 2017 (Figure 8.7.2.8). This also resulted in a higher estimated SSB for the period since 2012, although the magnitude of the difference was less than when the entire RFID data set was removed. For comparison, the same exercise was done removing the last year of data for the IESSNS (Figure 8.7.2.9). This resulted only in a minor (downward) revision of the recent estimates of SSB.

This shows that the RFID tagging data has a very strong weight on the assessment, and pulls recent estimates of abundance downward. This feature of the assessment has not been investigated in the previous years, although it was noted during the previous benchmark that the decision to include the RFID data resulted in a lower SSB in the recent years. WGWIDE recommends that this aspect of the assessment should be further studied, and that the better understanding of the relative weight of the different data sources should be gained. Since the tag recaptures are modelled with an error distribution (negative binomial) different from the error distribution used for the other observations (log normal), model parameters cannot be used to compare their relative weight.

### 8.7.3 State of the Stock

The stock summary is presented in Figure 8.7.3.1 and Table 8.7.3.1. The stock numbers-at-age and fishing mortality-at-age are presented in Tables 8.7.3.2-3, respectively. The spawning stock biomass is estimated to have increased almost continuously from just below 2 million tonnes in the late 1990s and early 2000s to 4.79 million tonnes in 2011 and subsequently declined continuously to reach a level just above 3 million tonnes in 2017. The fishing mortality has declined from levels close to $\mathrm{Flim}_{\lim }(0.46)$ in the mid-2000s to 0.26 in 2012 and has increased again since then to levels above $\mathrm{F}_{\mathrm{pa}}$. The recruitment time series from the assessment shows a clear increasing trend since the late 1990s with a succession of large year classes (2002, 2006, 2011 and 2014). The estimates for the year classes 2015 and 2016 indicate low recruitment, likely the lowest in the time series for the 2016 year class. There is insufficient information to estimate accurately the size of the 2017 year class. The estimate is very high but highly uncertain as it relies only on the age 0 catch data (in absence of a 2017 IBTS index).

There is some indication of changes in the selectivity of the fishery over the last 20 years (Figure 8.7.3.2.). In the year 1994, the fishery seems to have exerted a high fishing mortality on the fish 7 years and older. This changed gradually until 2000, when the fishing mortality on younger ages (3- to 6-years) increased compared to the older fish. In the following years, the selectivity pattern changed towards a lower fishing mortality on the age-classes younger than 6 years until 2008 when it changed again towards a higher selection on the fish age 3 to 6 .

### 8.7.4 Additional exploratory runs with different selection criteria for the tagging data

### 8.7.4.1 Accounting for the geographic area of recapture

Exploratory analyses presented in Section 8.6 .4 suggest that the tag recapture rates may change according to the area of recapture. Potential biological explanations are given in the Section 8.6.4. If such is that case, it may be appropriate to take account of these differences in the assessment model. This can be done by estimating a different post
release survival rate for the different areas considered in Section 8.6.4 (see Figure 8.6.4.5).

The RFID tag dataset structured by area is different from the one used in the update assessment: for each recapture year, there can be up to 4 data points (for the 4 areas) instead of one for each cohort in each release year. In order to assess the effect of using an area effect in the model, the model without area effect therefore had to be run first on the data set structured by area.

Model parameters were slightly changed when replacing the RFID tagging data by the data set structured by area (Figure 8.7.4.1.1). A small reduction of the observation variance for the egg survey and an increase for the IESSNS are observed. The overdispersion for the RFID tags decreases slightly, but the parameter is extremely badly defined (such that the parameter standard deviation could not be estimated). Including the area effect has only a minor effect on the parameters, and the problem with the high uncertainty on the overdispersion for the RFID tag remains. The problem with the estimation of the parameters was even more acute when the model was configured with separate overdispersion parameters for each geographical area (result not shown). The estimated survival rates show some differences between areas, with lower values for the area 3, average values for the areas 1 and 4 (similar to the parameter estimate without area effect) and high value for the area 2 (Figure 8.7.4.1.2). These values are consistent with the observations made in Section 8.6.4. Introducing an area specific survival rate resulted in smaller changes in the recent SSB and F ( $+7 \%$ and $-7 \%$ for 2018 SSB and Fbar respectively). Changing the RFID tagging dataset without any change in model configuration resulted in a downward revision by $-18 \%$ of the recent SSB estimates and an upward revision of the same magnitude in the fishing mortality (Figure 8.7.4.1.3).

Although the differences in model AIC and the differences in estimated survival rates between areas suggest that it might be appropriate to take account of recapture area in the model, the issues found with parameter estimation deserve further attention. It is likely that data series may still be too short for some areas where the scanning of the RFID tags started only in the recent years.

### 8.7.4.2 Influence of the number of years before recapture

Investigations presented in Section 8.6.4 suggest that the recapture rates of a cohort tagged in a given year tend to decrease with the number of years separating tagging and recapture. In the context of the assessment model, this could be translated in differences in survival rate with the number of years between release and recapture. Since the model assumes a unique rate, this would result in a pattern in the residuals, with larger values (for a given cohort tagged in a given years) for the first recapture years, and lower residuals for fish that remained longer in the sea.

In order to investigate the existence of such patterns, the residuals were grouped by year-class and release year (or equivalently age at release). For each group, the residuals for the different recapture years were then centred (subtracting the mean) and inspected for pattern. Figure 8.7.4.2.1 shows that for a number of instances the residuals tend to decrease with the number of years spent at sea, which supports the hypothesis that mortality increase with the number of years after tagging. However, this is not the case of each cohort/age at release combination.

The existence of such patterns may indicate that the model is not formulated appropriately, and that the cumulative mortality with the successive years spent before recapture should be explicitly accounted for in the model. As an attempt to remove this potential bias in the assessment, the model was run using only the recapture of the first

2 years after tagging. This assessment estimated a lower overdispersion of the RFID tags, and a $20 \%$ higher estimated survival rate (which was to be expected if indeed mortality due to tagging continues in the years after release). There was however a slight increase in the uncertainty around these parameters. The corresponding stock trajectories are substantially revised (by $+20 \%$ for 2017 SSB and by $-21 \%$ for Fbar) (Figure 8.7.4.2.2).

This issue deserves further investigations, and potential model modifications to better model mortality after release should be investigated in a future benchmark.

### 8.7.5 Quality of the assessment

## Parametric uncertainty

Large confidence intervals are associated with the SSB in the years before 1992 (Figure 8.7.3.1 and Figure 8.7.5.1). This results from the absence of information from the egg survey index, the downweighing of the information from the catches and the assessment being only driven by the tagging data and natural mortality in the early period. The confidence intervals become narrower from the early 1990s to the mid-2000s, corresponding to the period where information is available from the egg survey index, the tagging data and (partially) catches. The uncertainty increases slightly in the most recent years and the SSB estimate for 2017 is estimated with a precision of $+/-28 \%$ (Figure 8.7.3.1 and Table 8.7.3.1). There is generally also a corresponding large uncertainty on the fishing mortality, especially before 1995. The estimate of Fbar4-8 in 2017 has a precision of $+/-31 \%$. The uncertainty on the recruitment is high for the years before 1998 (precision of on average $+/-63 \%$ ). The precision improves for the years for which the recruitment index is available (+/-39\%) except for the last estimated recruitment (+/99\%).

## Model instability

The retrospective analysis was carried out for 5 retro years, by fitting the assessment using the 2018 data, removing successively 1 year of data (Figure 8.7.5.2). Since some of the time series are still short (8 years for the IESSNS index, 6 years for the RFID tags), the parameters corresponding to these sources of data are expected to change from year to year, until the time series are long enough to have stable estimates.

There is no strong retrospective pattern observed in the SSB, as indicated by the reasonably low Mohn's rho value (i.e. average relative bias of retrospective estimates; Mohn, 1999; Brooks and Legault, 2016). All runs, except the one removing 5 years of data, provide estimates which are within the confidence intervals of the current assessment. Differences in the estimated $F_{b a r}$ values are larger than for SSB and tend to show a pattern to towards systematic overestimation, as indicated by the Mohn's rho value of 0.23 . Recruitment appears to be quite consistently estimated.

## Model behaviour

The realisation of the process error in the model was also inspected. The process error expressed as annual deviations in abundances-at-age (Figure 8.7.5.3) shows indications of some pattern across time and ages. There is a predominance of positive deviations in the recent years for age-classes 4 to 8 . While process error is assumed to be independent and identically distributed, there is clear evidence of correlations in the realisation of the process error in the mackerel assessment, which appears to be correlated both across age-classes and temporarily.

The temporal autocorrelation can also be visualised if the process error is expressed in term of biomass (process error expressed as deviations in abundances-at-age multiplied by weight at age and summed over all age classes, Figure 8.7.5.4). Periods with positive values (when the model globally estimates larger abundances-at-age than corresponding to the survival equation) have been alternating with periods with negative values (between 2004 and 2007). For the years since 2010 the cumulated process error remains positive, with the magnitude reaching a third of the volume of the catches for 2009. The reason for this misbehaviour of the model could not be identified. It should be noted, however, that the magnitude and autocorrelation of the biomass cumulated process error in the 2018 assessment is lower than in the previous year's assessment.

### 8.8 Short term forecast

The short-term forecast provides estimates of SSB and catch in 2019 and 2020, given assumption of the current year's (also called intermediate year) catch and a range of management options for the catch in 2019.

All procedures used this year follow those used in the benchmark of 2014 as described in the Stock Annex.

### 8.8.1 Intermediate year catch estimation

Estimation of catch in the intermediate year (2018) is based on declared quotas and interannual transfers as shown in the text table in Section 8.1.

### 8.8.2 Initial abundances at age

The recruitment estimate at age 0 from the assessment in the terminal assessment year (2017) was considered too uncertain to be used, because this year class has not yet fully recruited into the fishery. The last recruitment estimate is normally replaced by predictions from the RCT3 software (Shepherd, 1997). The RCT3 software evaluates the historical performance of the IBTS recruitment index, by performing a linear regression between the index and the SAM estimates over the period 1998 to the year before the terminal year. The recruitment is then calculated as a weighted mean of the prediction from this linear regression based on the IBTS index value, and a time tapered geometric mean of the SAM estimates from 1990 to the year before the terminal year. The time tapered geometric mean gives the latest years more weight than a geometric mean. This is done because the recent productivity of the stock appears different than in the 1990's.

The recruitment index for 2016 and 2017 could not be calculated (see Section 8.6.2). The time tapered geometric mean ( 5267776 ) from 1990-2015 was therefore used as the recruitment in 2016 and 2017 in the forecast. This is equivalent to the standard method using RCT3, except that (missing) recruit index value has no influence.

### 8.8.3 Short term forecast

A deterministic short-term forecast was calculated using FLR. Table 8.8.3.1 lists the input data and Tables 8.8.3.2 and 8.8.3.3 provide projections for various fishing mortality multipliers and catch constraints in 2019.
Assuming catches for 2018 of 1001 kt , F was estimated at 0.46 (close to Flim) and SSB at 2.35 Mt (below $\mathrm{B}_{\mathrm{pa}}$ ) in spring 2018. If catches in 2019 equal the catch in 2018, F is expected to increase to 0.66 (above Flim) in 2019 with a corresponding reduction in SSB to
1.98 Mt in spring 2019, which is close to $B_{\lim }(1.94 \mathrm{Mt})$. Assuming an F of 0.66 again in 2020, the SSB will decrease further to 1.71 Mt in spring 2020.
Following the MSY approach, exploitation in 2019 shall be at FMSY * SSB(2019) /MSY $B_{\text {trigger, }}$ because SSB is predicted to be below MSY $B_{\text {trigger }}(2.57 \mathrm{Mt})$ in spring 2019. This is equivalent to an F at 0.173 , catches of 318 kt and a reduction in SSB to 2.12 Mt in spring 2019 (-10 \% change). This is still below Btrigger. During the subsequent year, SSB is predicted increase with $10 \%$ to 2.33 Mt in spring 2020.

### 8.9 Biological Reference Points

A long-term management plan evaluation was conducted in 2017 (ICES, 2017b) which resulted in the adoption of new reference points for NEA mackerel stock by ICES.

### 8.9.1 Precautionary reference points

$B_{\text {lim }}$ - There is no evidence of significant reduction in recruitment at low SSB within the time series hence the previous basis for $B_{\lim }$ was retained. Blim is taken as Bloss, the lowest estimate of spawning stock biomass from the revised assessment. This was estimated to have occurred in 2002; Bloss $=1940000 \mathrm{t}$.
$F_{l i m}$ - Flim is derived from $B_{\lim }$ and is determined from the long-term equilibrium simulations as the F that on average would bring the stock to $\mathrm{Blim}_{\mathrm{lim}} ; \mathrm{F}_{\mathrm{lim}}=0.48$.
$\boldsymbol{B}_{p a}$ - The ICES basis for advice requires that a precautionary safety margin incorporating the uncertainty in actual stock estimates leads to a precautionary reference point $\mathrm{B}_{\mathrm{pa}}$, which is a biomass reference point with a high probability of being above $\mathrm{B}_{\mathrm{lim}} . \mathrm{B}_{\mathrm{pa}}$ was calculated as $B_{\text {lim }} \cdot \exp (1.645 \cdot \sigma)$ where $\sigma=0.17$ (the estimate of uncertainty associated with spawning biomass in the terminal year in the assessment, 2016, as estimated in the 2017 management plan evaluation); $\mathrm{B}_{\mathrm{pa}}=2570000 \mathrm{t}$.
$\boldsymbol{F}_{p a}$-The ICES basis for advice requires that a precautionary safety margin incorporating the uncertainty in actual stock estimates leads to a precautionary reference point $\mathrm{F}_{\mathrm{pa}}$, which is a fishing mortality reference point designed to avoid reaching Flim. Consequently, $\mathrm{F}_{\mathrm{pa}}$ was calculated as $\mathrm{F}_{\mathrm{lim}} * \exp (1.645 \sigma)$ where $\sigma=0.20$ default value was taken following the guidelines, as the estimated standard deviation of $\ln (\mathrm{F})$ in the final assessment year (2016) provided by the SAM assessment (i.e. $\sigma=0.14$ corresponding to the uncertainty of $\ln \left(\mathrm{F}_{2015}\right)$ ) was smaller than 0.20 but considered unrealistically low.; $\mathrm{F}_{\mathrm{pa}}=0.35$.

### 8.9.2 MSY reference points

The ICES MSY framework specifies a target fishing mortality, FMSY, which, over the long term, maximises yield, and also a spawning biomass, MSY Btrigger, below which target fishing mortality is reduced linearly relative to the SSB B trigger ratio.

Following the ICES guidelines (ICES, 2013b), long term equilibrium simulations indicated that $\mathrm{F}=0.21$ would be an appropriate $\mathrm{F}_{\text {msy }}$ target as on average it resulted in the highest mean yields in the long term, with a low probability (less than $5 \%$ ) of reducing the spawning biomass below Blim.

The ICES basis for advice notes that, in general, Fmsy should be lower than $\mathrm{F}_{\mathrm{pa}}$, and MSY $B_{\text {trigger }}$ should be equal to or higher than $B_{\text {pa }}$. Simulations indicated that potential values for MSY Btrigger were below $\mathrm{B}_{\mathrm{pa}}$. Following the ICES procedure MSY Btrigger was set equal to $\mathrm{Bpa}_{\mathrm{pa}} 2570000 \mathrm{t}$.

Updated ICES reference points for NEA mackerel

| Type |  | Value | Technical basis |
| :---: | :---: | :---: | :---: |
| MSY <br> approach | MSY Btrigger | 2.57 million tonnes | $\mathrm{Bpa}^{1}$ |
|  | FMSY | 0.21 | Stochastic simulations ${ }^{1}$ |
| Precautionary approach | Blim | 1.94 million tonnes | Bloss in $2002{ }^{2}$ |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 2.57 million tonnes | $\mathrm{B}_{\lim } \times \exp (1.654 \times \sigma), \sigma=0.17{ }^{1}$ |
|  | Flim | 0.48 | F that on average leads to $\mathrm{Blim}^{1}{ }^{1}$ |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.35 | $\mathrm{Flim} \times \exp (1.654 \times \sigma), \sigma=0.20^{1}$ |

${ }^{1} 2017$ management plan evaluation (ICES, 2017b)
${ }^{2} 2017$ benchmark assessment (ICES, 2017a)

### 8.10 Comparison with previous assessment and forecast

## Assessment

The last available assessment used for providing advice was carried out in 2017 at WGWIDE. The new 2018 WGWIDE assessment gives a slightly different perception of the recent development of the stock (Figure 8.10.1). While the previous assessment gave the perception of a stock stable at high levels after 2011, the new assessment now indicates that the stock has been declining since 2011. Conversely the new assessment suggests that F has been increasing constantly since 2011, while the previous assessment indicated a less pronounced increase.
The differences in the 2016 TSB and SSB estimates between the previous and the present assessments are moderate, of $-11 \%$ in both cases. The upward revision of the 2016 fishing mortality estimate is small, of $4 \%$.

|  | TSB 2016 | SSB 2016 | F4-8 2016 |
| :--- | :--- | :--- | :--- |
| Values |  |  |  |
| 2017 WGWIDE | 4752576 | 3970992 | 0.322 |
| 2018 WGWIDE | 4216702 | 3527235 | 0.335 |
| $\%$ difference | $-11 \%$ | $-11 \%$ | $+4 \%$ |

The exploratory runs presented in Section 8.7. 2 showed that removing the last year of tagging data (recaptures from 2017) modified strongly the perception of the stock. The estimated SSB is in this case more similar to last year's assessment (see Figure 8.7.2.8). The same section shows that the 2018 IESSNS data point has little influence on the recent SSB and Fbar estimates (Figure 8.7.2.9). The recaptures from 2017 added in this update assessment inform the model on the abundance-at-age for ages 2 to 12 for the period 2011 to 2017 (so basically 1 additional year of RFID data may potentially provide as much information as the entire IESSNS index).

Inspecting the changes in the estimated model parameters can help understand the reason for these revisions (Figure 8.10.2). The addition of an additional year of data has slightly modified the relative weight of the different data sources: the estimated observation standard deviation has decreased for the catches and the egg survey, and increased for the IESSNS age 4-11 and the recruitment index. The overdispersion for the

RFID tags also increased. The model also estimates this year more variable recruitment and fishing mortality, and a smaller process error.

The uncertainty on the parameter estimates has decreased for a number of parameters (Figure 8.10.2). It is for instance the case for the observation standard deviations for the IESSNS, the overdispersion of the RFID tags, and some of the catchabilities estimates. However, the observation standard deviation for the catches has become slightly more uncertain. The joint uncertainty on recent SSB and Fart-8 in this year's assessment is lower than for last year's assessment (Figure 8.10.3).

## Forecast

The prediction of the mackerel catch for 2017 used for the short-term forecast in the advice given last year was very close to the actual 2017 catch reported in 2018 and used in the present assessment (text table below). The new assessment produced an estimate of the SSB in 2017 10.5\% lower than the 2017 forecast prediction. The fishing mortality Fbar-8 for 2017 estimated this year is $6.2 \%$ lower than the value estimated by the short term forecast in the previous assessment. Most of these discrepancies can be explained by the revision of the perception of the stock described above.

|  | Catch (2017) | SSB (2017) | F4-8(2017) |
| :--- | :--- | :--- | :--- |
| 2017 WGWIDE forecast | 1178850 t | 3443926 t | 0.405 |
| 2018 WGWIDE assessment | 1155944 t | 3081442 t | 0.38 |
| \% difference | $-1.9 \%$ | $-10.5 \%$ | $-6.2 \%$ |

### 8.11 Management Considerations

Details and discussion on quality issues in this year's assessment is given in Section 8.7 above.

The Atlantic mackerel in the Northeast Atlantic is traditionally characterised as three distinct 'spawning components': the southern component, the western component and the North Sea component. The basis for the components is derived from tagging experiments (ICES, 1974), however, the methods normally used to identify stocks or components (e.g. ectoparasite infections, blood phenotypes, otolith shapes and genetics) have not been able to demonstrate significant differences between animals from different components. The mackerel in the Northeast Atlantic appears on one hand to mix extensively whilst, on the other hand, exhibit some tendency for homing (Jansen et al., 2013; Jansen and Gislason, 2013). Consequently, it cannot be considered either a panmictic population, nor a population that is composed of isolated components (Jansen and Gislason, 2013).

Nevertheless, stock components are still being used to identify the different spawning areas where mackerel are known to spawn. The trends in the different components is derived from the triennial egg survey in the western and southern area and a dedicated egg survey in the North Sea the year following the western survey.

Since the mid-1970s, ICES has continuously recommended conservation measures for the North Sea component of the Northeast Atlantic mackerel stock (e.g. ICES, 1974; ICES, 1981). The measures advised by ICES to protect the North Sea spawning component (i.e. closed areas and minimum landing size) aimed to promote the conditions that make a recovery of this component possible.

The recommended closure of Division 4.a for fishing during the first half of the year is based on the perception that the western mackerel enter the North Sea in July/August, and remain there until December before migrating to their spawning areas. Updated observations from the late 1990s suggested that this return migration actually started in mid- to late February (Jansen et al., 2012). The EU TAC regulations stated that within the limits of the quota for the western component (ICES Subareas and Divisions 6, 7, $8 . a, b, d, e, 5 b(E U), 2 a(n o n-E U), 12,14)$, a certain quantity of this stock may be caught in 4.a during the periods 1 January to 15 February and 1 September to 31 December. Up to $2010,30 \%$ of the Western EU TAC of mackerel (MAC/2CX14-) could be taken in 4.a. From 2011 onwards, this percentage has been set at $40 \%$ and from 2015 at $60 \%$.

The minimum landing size (MLS) for mackerel is currently set at 30 cm for the North Sea and 20 cm in the western area. The historical basis for the setting of minimum landing sizes is described in a working document to WGWIDE in 2015 (Pastoors, 2015). The MLS of 30 cm in the North Sea was originally introduced by Norway in 1971 and was intended to protect the very strong 1969 year class from exploitation in the industrial fishery. The 30 cm later became the norm for the North Sea MLS while the MLS for mackerel in western waters was set at 20 cm . In the early 1990s, ICES recommended that, because of mixing of juvenile and adult mackerel on western waters fishing grounds, the adoption of a 30 cm minimum landing size for mackerel was not desirable as it could lead to increased discarding (ICES, 1990; 1991). A substantial part of the catch of (western) NEA mackerel is taken in ICES division 4.a during the period October until mid-February to which the 30 cm MLS applies even though there is limited understanding on the effectiveness of minimum landing sizes in achieving certain conservation benefits (STECF, 2015).

### 8.12 Ecosystem considerations

An overview of the main ecosystem drivers possibly affecting the different life-stages of Northeast Atlantic mackerel and relevant observations are given in the Stock Annex. The discussion here is limited to recent features of relevance.

## Production (recruitment and growth)

Mackerel recruitment (age 1) has been higher since 2001 compared to previous decades with several very large cohorts (Jansen, 2016). Increasing stock size was suggested to have an effect through density driven expansion of the spawning area into new areas with Calanus in oceanic areas west of the North European continental shelf (Jansen, 2016). There are several indications of a shift in spawning and mackerel recruitment/larvae and juvenile areas towards northern and north-eastern areas preceding the 2016 mackerel spawning (ICES, 2016b; Nøttestad et al. 2018). This northerly shift in spawning and recruitment pattern of NEA mackerel seem to have continued also in 2017 and 2018 (Nøttestad et al., 2018). The incoming 2017-year class has the largest age1 index value recorded in IESSNS and is $150 \%$ larger than the incoming age- 1 cohort in 2017 (ICES, 2018a).

During the recent decade, mackerel length- and weight-at-age declined substantially for all ages (Jansen and Burns, 2015; Ólafsdóttir et al., 2015). Growth of 0-3 years old mackerel decreased from 1998 to 2012. Mean length at age 0 decreased by 3.6 cm , however the growth differed substantially among cohorts (Jansen and Burns, 2015). For the 3-8 years old mackerel, the average size was reduced by 3.7 cm and 175 g from 2002 to 2013 (Ólafsdóttir et al., 2015). The variations in growth of mackerel in all ages are correlated with mackerel density. Furthermore, the density dependent regulation of growth from younger juveniles to older adult mackerel, appears to reflect the spatial
dynamics observed in the migration patterns during the feeding season. (Jansen and Burns, 2015; Ólafsdóttir et al., 2015). Growth rates of the juveniles were tightly correlated with the density of juveniles in the nursery areas (Jansen and Burns, 2015). For adult mackerel (age 3-8) growth rates were correlated with the combined effects of mackerel and herring stock sizes (Ólafsdóttir et al., 2015). Conspecific density-dependence was most likely mediated via intensified competition associated with greater mackerel density.

Furthermore, the last few years after 2014, the recruitment appear weaker for NEA mackerel (ICES, 2017b; 2018c) and the density dependent growth has stabilized and mean weights per age group have even slightly increased during the last 2-3 years for several age groups (ICES, 2018c).

## Spatial mackerel distribution and timing

In the mid-2000s, summer feeding distribution of Northeast Atlantic mackerel (Scomber scombrus) in Nordic Seas began expanding into new areas (Nøttestad et al., 2016). During 2007-2016 period mackerel distribution range increased three-fold and the centre-of-gravity shifted westward by 1650 km and northward by 400 km . Distribution range peaked in 2014 and was positively correlated to Spawning Stock Biomass (SSB).

After a mackerel stock expansion during the feeding season in summer from 1.3 million $\mathrm{km}^{2}$ in 2007 to at least 2.9 million $\mathrm{km}^{2}$ in 2014, mainly towards western and northern regions of the Nordic seas (Nøttestad et al., 2016c), we have now a slightly decreased distribution area of mackerel in the Nordic Seas (Nøttestad et al. 2017; ICES, 2018a). The survey coverage area was 2.8 million square kilometres in 2018, which is the same as in 2017 (Nøttestad et al. 2017; ICES, 2018a). The mackerel appeared more evenly distributed within the survey area and more easterly distributed in 2018 than in 2017 (ICES, 2018a). This difference in distribution primarily consists of a marked biomass decline of $76 \%$ in the west. In the eastern areas, the decline was less with $21 \%$. Furthermore, there was also an eastward shift of distribution and centre of gravity within the Norwegian Sea (ICES, 2018a).
Geographical distribution of the 2016 cohort at age 0 and 1 was different from the traditional juvenile distribution patterns. The 2016 cohort was observed from latitude 60$71^{\circ} \mathrm{N}$ along the coast and offshore areas of Norway based on various survey data and fishing data (Nøttestad et al., 2018). Traditional, 0- and 1-group of mackerel reside further south in waters of the southernmost part of Norway.
An historical and very pronounced shift in distribution of juvenile mackerel took place along the Norwegian coast starting off during the autumn of 2016 onwards (ICES, 2017b; Nøttestad et al., 2017; Nøttestad et al., 2018). This also coincided with increased number of adult and mature mackerel in northern waters from May to July 2016 (ICES, 2016) as well as from May to July 2017-2018 (ICES, 2017; Nøttestad et al. 2018). The prevalence of adult mackerel in the northern North Sea and southern Norwegian Sea increased markedly in first quarter and second quarter 2016, compared to the two previous years in 2014 and 2015, suggesting a shift in spawning of mackerel towards the north and northeast (Nøttestad et al., 2018).

The results showed also a marked increase in the presence of zero-year and one-year old mackerel in the northern North Sea and Skagerrak first quarter 2017, compared to first quarter 2014-2016. In the second quarter there were strong indications of spawning mackerel outside and north of the spatial and temporal coverage during the 2016 mackerel egg survey (Nøttestad et al., 2018).

## Spatial mackerel distribution related to environmental conditions

Mackerel was present in temperatures ranging from $5{ }^{\circ} \mathrm{C}$ to $15^{\circ} \mathrm{C}$, but preferred areas with temperatures between $9{ }^{\circ} \mathrm{C}$ and $13{ }^{\circ} \mathrm{C}$ according to univariate quotient analysis according to Ólafsdóttir et al. (2018). Generalized additive models showed that both mackerel occurrence and density were positively related to location, ambient temperature, meso-zooplankton density and SSB, explaining $47 \%$ and $32 \%$ of deviance, respectively (Ólafsdóttir et al. 2018). Mackerel relative mean weight-at-length was positively related to location, day-of-year, temperature and SSB, but not with mesozooplankton density, explaining $40 \%$ of the deviance. Geographical expansion of mackerel during the summer feeding season in Nordic Seas was driven by increasing mackerel stock size and constrained by availability of preferred temperature and abundance of meso-zooplankton (Ólafsdóttir et al., 2018). Marine climate with multidecadal variability probably impacted the observed distributional changes but were not evaluated. Our results were limited to the direct effects of temperature, meso-zooplankton abundance, and SSB on distribution range during the last two decades (1997-2016) and should be viewed as such. In the 2018 IESSNS a marked change in the spatial distribution of mackerel was observed with lower densities of mackerel in the western distributions areas (East Greenland and Iceland) as compared to the recent years (see Figure 8.6.3.1). It is not clear what causes this distributional shift, but the SST were $1-1.5^{\circ} \mathrm{C}$ lower in the western and south-western areas as compared to a 20 years mean (19992009) might partly explain such changes (ICES, 2018a).

## Trophic interactions

There are strong indications for interspecific competition for food between NSS-herring, blue whiting and mackerel (Huse et al., 2012). According to Langøy et al. (2012), Debes et al. (2012), Óskarsson et al. (2015) and Bachiller et al. (2016), the herring may suffer from this competition, as mackerel had higher stomach fullness index than herring and the herring stomach composition is different from previous periods when mackerel stock size was smaller. Langøy et al (2012) and Debes et al. (2012) also found that mackerel consumed wider range of prey species than herring. Mackerel may thus be thriving better in periods with low zooplankton abundances. Feeding incidence increased with decreasing temperature as well as stomach filling degree, indicating that feeding activity is highest in areas associated with colder water masses (Bachiller et al., 2016). A bioenergetics model being developed by Bachiller et al. (2016) estimates that the NEA mackerel, NSS herring and blue whiting can consume between 122 and 135 million tonnes of zooplankton per year (2005-2010) This is higher than that estimated in previous studies (e.g. Utne et al., 2012; Skjoldal et al., 2004). NEA mackerel feeding rate can consequently be as high as that of the NSS herring in some years. Geographical distribution overlap between mackerel and NSS herring during the summer feeding season is highest in the south-western part of the Norwegian Sea (Faroe and east Icelandic area) (Nøttestad et al. 2016; 2017; Ólafsdóttir et al., 2018). The spatio-temporal overlap between mackerel and herring was highest in the southern and south-western part of the Norwegian Sea in 2018 (ICES, 2018a). This is similar as seen in previous years (Nøttestad et al. 2016; 2017). There was practically no overlap between NEA mackerel and NSSH in the central and northern part of the Norwegian Sea in 2018, mainly because of very limited amounts of herring in this area (ICES, 2018a).

The increase of 0 - and 1-groups of NEA mackerel found along major coastlines of Norway both in 2016 and 2017 (Nøttestad et al., 2018), has created some interesting new trophic interactions. Increasingly numbers of adult Atlantic bluefin tuna (Thynnus thunnus), with an average size of approximately 200 kg , have been documented to feed
on 0-group mackerel from the 2016, 2017-year classes during the commercial bluefin tuna fishery in Norway (Nøttestad et al., 2017b). Additionally, the new situation of numerous 0- and 1-group mackerel in Norwegian coastal waters have created favourable feeding possibilities for larger cod, saithe, marine mammals and seabirds in these waters. Repeated stomach samples from several species document that smaller sized mackerel is now eaten by different predators in northern waters $\left(60-70^{\circ} \mathrm{N}\right)$.

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### 8.14 Tables

Table 8.2.1. 2017 Mackerel fleet composition of major mackerel catching nations.

| Country | LEN (M) | ENGINE POWER (HP) | Gear | Storage |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 57-88 | 4077-8158 | Trawl | Tank | 8 |
| Faroe Islands | 50-69 | $3460-8000 \mathrm{kw}$ | Purse Seine/Trawl | RSW | 3 |
|  | 70-76 | $3920-7500 \mathrm{kw}$ | Purse Seine/Trawl | RSW | 4 |
|  | 73-104 | $6000-6600 \mathrm{kw}$ | Trawl | Freezer | 2 |
|  | 15-49 | $300-1940 \mathrm{kw}$ | Trawl |  | 20 |
|  | 50-79 | $3000-7680 \mathrm{kw}$ | Trawl |  | 7 |
| France | <24 |  | Trawl |  | 1230 |
|  | >24 |  | Trawl |  | 36 |
| Germany | 90-140 | 3800-12000 | Single Midwater Trawl | Freezer | 4 |
| Greenland | 66-80 | 4011-10034 | Trawl | RSW | 9 |
|  | 55-88 | 3712-8164 | Trawl | Freezer/RSW | 5 |
|  | 65-120 | 3002-9517 | Trawl | Freezer | 12 |
| Iceland | 51-60 | 2502-4079 | Single Midwater Trawl | RSW, Freezer | 6 |
|  | 61-70 | 2000-7507 | Single Midwater Trawl | RSW, Freezer | 17 |
|  | 71-80 | 3200-11257 | Single Midwater Trawl | RSW, Freezer | 12 |
|  | >80 | 8051 | Single Midwater Trawl | Freezer | 1 |
| Ireland | 27m-65m | 522-2720 | Pair Midwater Trawl | RSW | 14 |
|  | 14m-45m | 160-1119 | Pair Midwater Trawl | Dryhold | 23 |
|  | $51 \mathrm{~m}-71 \mathrm{~m}$ | 1007-3840 | Midwater Trawl | RSW | 8 |
|  | $12 \mathrm{~m}-17 \mathrm{~m}$ | 90-171 | Midwater Trawl | Dryhold | 2 |
| Netherlands | 55 | 2125 | Pair Midwater Trawl | Freezer | 1 |
|  | 88-145 | 4400-10455 | Single Midwater Trawl | Freezer | 9 |
| Norway | $60-85 \mathrm{~m}$ |  | Purse seiner | RSW | 78 |
|  | $30-40 \mathrm{~m}$ |  | Purse seiner | Dryhold, RSW | 16 |
|  | $10-17 \mathrm{~m}$ |  | Purse seiner | Dryhold | 178 |
|  | $10-17 \mathrm{~m}$ |  | Hook and line/nets | Dryhold | 169 |
|  | $10-17 \mathrm{~m}$ |  | PS/hooks/nets | Dryhold | 200 |
|  | $30-40 \mathrm{~m}$ |  | Trawl | Dryhold.Tankhold | 17 |
| Portugal | 0-10 |  | Other |  | 94 |
|  | 10-20 |  | OTB |  | 3 |
|  | 10-20 |  | Other |  | 86 |
|  | 20-30 |  | OTB |  | 27 |
|  | 20-30 |  | Other |  | 16 |
|  | 30-40 |  | Trawl |  | 7 |
| Spain | 12-18 | 80-294 | Trawl | Dryhold | 12 |
|  | 18-24 | 96-344 | Trawl | Dryhold | 30 |
|  | 24-40 | 191-876 | Trawl | Dryhold | 72 |
|  | 40- | 353 | Trawl | Dryhold | 2 |
|  | 0-10 | 34-44 | Purse Seine | Dryhold | 2 |
|  | 10-12 | 20-106 | Purse Seine | Dryhold | 13 |
|  | 12-18 | 21-245 | Purse Seine | Dryhold | 112 |
|  | 18-24 | 70-397 | Purse Seine | Dryhold | 100 |
|  | 24-40 | 140-809 | Purse Seine | Dryhold | 99 |
|  | 0-10 | 3-74 | Artisanal | Dryhold | 329 |
|  | 10-12 | 12-118 | Artisanal | Dryhold | 203 |
|  | 12-18 | 18-239 | Artisanal | Dryhold | 208 |
|  | 18-24 | 59-368 | Artisanal | Dryhold | 40 |
|  | 24-40 | 129-368 | Artisanal | Dryhold | 11 |
| 1 RSW = refrig | seawater. |  |  |  |  |

Table 8.2.4.1. Overview of major existing regulations on mackerel catches.

| Technical measure | National/International level | Specification | Note |
| :---: | :---: | :---: | :---: |
| Catch limitation | Coastal States/NEAFC | 2010-2018 | Not agreed |
| Management strategy (EU, NO, FO agreement London 12. Oct. 2014) | European (EU, NO, FO) | If SSB >= 3.000.000t, F $=0.24$ <br> If SSB is less than $3.000 .000 \mathrm{t}, \mathrm{F}=$ $0.24 \quad * \quad$ SSB/3.000.000 TAC should not be changed more than 20\% A party may transfer up to $10 \%$ of unutilised quota to the next year | ot agreed by all arties |
| Management strategy with updated reference points 2017 (EU, NO, FO agreement London 11. Oct. 2017) | European (EU, NO, FO) | If SSB $>=2.570 .000 \mathrm{t}, \mathrm{F}=0.21$ <br> If SSB is less than $2.570 .000 \mathrm{t}, \mathrm{F}=$ $0.21 \quad * \quad$ SSB/2.570.000 <br> TAC should not be changed more than $+25 \%$ or $-20 \%$ A party may transfer up to $10 \%$ of unutilised quota to the next year A party may fish up to $10 \%$ beyond the allocated quota, that have to be deduced from next years quota. | Not agreed by all parties |
| Minimum (North Sea) $\quad$ size | European (EU, NO, FO) | 30 cm in the North Sea |  |
| ```Minimum size (all areas except North Sea)``` | European (EU, NO) | 20 cm in all areas except North Sea | $10 \%$ undersized allowed |
| Minimum size | National (NO) | 30 cm in all areas |  |
| Catch limitation | European (EU, NO, FO) | Within the limits of the quota for the western component (VI,VII, VIIIabde, $\mathrm{Vb}(\mathrm{EC})$, IIa(nonEC), XII, XIV), a certain quantity may be taken from IVa but only during the periods 1 January to 15 February and 1 October to 31 December. |  |
| Area closure | National (UK) | South-West Mackerel Box off Cornwall | Except where the weight of the mackerel does not exceed $15 \%$ by liveweight of the total quantities of mackerel and other marine organisms onboard which have been caught in this area |
| Area limitations | National (IS) | Pelagic trawl fishery only allowed outside of 200 m depth contours around Iceland and/or 12 nm from the coast. |  |
| National catch <br> limitations by gear, <br> semester and area  | National (ES) | $28.74 \%$ of the Spanish national quota is assigned for the trawl fishery, $34.29 \%$ for purse seiners and $36.97 \%$ for the artisanal fishery | Since 2015, the trawl fishery has the individual quotas assigned by vessel. |
| Discard prohibition | National (NO, IS, FO) | All discarding is prohibited for Norwegian, Icelandic and Faroese vessels |  |
| Landing Obligation | European | From 2015 onwards a landing obligation for European Union fisheries is in place for small pelagics including mackerel, horse mackerel, blue whiting and herring. <br> Since 2016 is also partly in place for demersal fisheries. |  |

Table 8.4.1.1. NE Atlantic Mackerel. ICES estimated catches by area (t). Discards not estimated prior to 1978 (data submitted by Working Group members).

| $\begin{gathered} \text { YEA } \\ R \end{gathered}$ | $\begin{gathered} \text { Subare } \\ \text { A } 6 \end{gathered}$ | Subarea 7 and Divisions 8.abde |  |  | Subareas 3 <br> AND 4 |  |  | Subareas 125 AND 14 |  |  | Divisions 8.CAND 9.A |  |  | Total |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Di sc | Catch | Ldg | $\begin{aligned} & \mathrm{Di} \\ & \mathrm{sc} \end{aligned}$ | Catch | Ldg | Dis c | Catc h | Ldg | Disc | Catch |
| $\begin{aligned} & 196 \\ & 9 \end{aligned}$ | 4800 |  | 4800 | 47404 |  | 47404 | 739175 |  | $\begin{aligned} & 73917 \\ & 5 \end{aligned}$ | 7 |  | 7 | 42526 |  | $\begin{aligned} & 4252 \\ & 6 \end{aligned}$ | $\begin{aligned} & 83391 \\ & 2 \end{aligned}$ |  | $\begin{aligned} & 83391 \\ & 2 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 0 \end{aligned}$ | 3900 |  | 3900 | 72822 |  | 72822 | 322451 |  | $\begin{aligned} & 32245 \\ & 1 \end{aligned}$ | 163 |  | 163 | 70172 |  | $\begin{aligned} & 7017 \\ & 2 \end{aligned}$ | $\begin{aligned} & 46950 \\ & 8 \end{aligned}$ |  | $\begin{aligned} & 46950 \\ & 8 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 1 \end{aligned}$ | 10200 |  | 10200 | 89745 |  | 89745 | 243673 |  | $\begin{aligned} & 24367 \\ & 3 \end{aligned}$ | 358 |  | 358 | 32942 |  | $\begin{aligned} & 3294 \\ & 2 \end{aligned}$ | $\begin{aligned} & 37691 \\ & 8 \end{aligned}$ |  | $\begin{aligned} & 37691 \\ & 8 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 2 \end{aligned}$ | 13000 |  | 13000 | $\begin{aligned} & 13028 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & 13028 \\ & 0 \end{aligned}$ | 188599 |  | $\begin{aligned} & 18859 \\ & 9 \end{aligned}$ | 88 |  | 88 | 29262 |  | $\begin{aligned} & 2926 \\ & 2 \end{aligned}$ | $\begin{aligned} & 36122 \\ & 9 \end{aligned}$ |  | $\begin{aligned} & 36122 \\ & 9 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 3 \end{aligned}$ | 52200 |  | 52200 | $\begin{aligned} & 14480 \\ & 7 \end{aligned}$ |  | $\begin{aligned} & 14480 \\ & 7 \end{aligned}$ | 326519 |  | $\begin{aligned} & 32651 \\ & 9 \end{aligned}$ | 21600 |  | 21600 | 25967 |  | $\begin{aligned} & 2596 \\ & 7 \end{aligned}$ | $\begin{aligned} & 57109 \\ & 3 \end{aligned}$ |  | $\begin{aligned} & 57109 \\ & 3 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 4 \end{aligned}$ | 64100 |  | 64100 | $\begin{aligned} & 20766 \\ & 5 \end{aligned}$ |  | $\begin{aligned} & 20766 \\ & 5 \end{aligned}$ | 298391 |  | $\begin{aligned} & 29839 \\ & 1 \end{aligned}$ | 6800 |  | 6800 | 30630 |  | $\begin{aligned} & 3063 \\ & 0 \end{aligned}$ | $\begin{aligned} & 60758 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 60758 \\ & 6 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 5 \end{aligned}$ | 64800 |  | 64800 | $\begin{aligned} & 39599 \\ & 5 \end{aligned}$ |  | $\begin{aligned} & 39599 \\ & 5 \end{aligned}$ | 263062 |  | $\begin{aligned} & 26306 \\ & 2 \end{aligned}$ | 34700 |  | 34700 | 25457 |  | $\begin{aligned} & 2545 \\ & 7 \end{aligned}$ | $\begin{aligned} & 78401 \\ & 4 \\ & \hline \end{aligned}$ |  | $\begin{aligned} & 78401 \\ & 4 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 6 \end{aligned}$ | 67800 |  | 67800 | $\begin{aligned} & 42092 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & 42092 \\ & 0 \end{aligned}$ | 305709 |  | $\begin{aligned} & 30570 \\ & 9 \end{aligned}$ | 10500 |  | 10500 | 23306 |  | $\begin{aligned} & 2330 \\ & 6 \end{aligned}$ | $\begin{aligned} & 82823 \\ & 5 \end{aligned}$ |  | $\begin{aligned} & 82823 \\ & 5 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 7 \end{aligned}$ | 74800 |  | 74800 | $\begin{aligned} & 25910 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & 25910 \\ & 0 \end{aligned}$ | 259531 |  | $\begin{aligned} & 25953 \\ & 1 \end{aligned}$ | 1400 |  | 1400 | 25416 |  | $\begin{aligned} & 2541 \\ & 6 \end{aligned}$ | $\begin{aligned} & 62024 \\ & 7 \end{aligned}$ |  | $\begin{aligned} & 62024 \\ & 7 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 8 \end{aligned}$ | 151700 | $\begin{aligned} & 1510 \\ & 0 \end{aligned}$ | $\begin{aligned} & 16680 \\ & 0 \end{aligned}$ | $\begin{aligned} & 35550 \\ & 0 \end{aligned}$ | $\begin{aligned} & 3550 \\ & 0 \end{aligned}$ | $\begin{aligned} & 39100 \\ & 0 \end{aligned}$ | 148817 |  | $\begin{aligned} & 14881 \\ & 7 \\ & \hline \end{aligned}$ | 4200 |  | 4200 | 25909 |  | $\begin{aligned} & 2590 \\ & 9 \end{aligned}$ | $\begin{aligned} & 68612 \\ & 6 \end{aligned}$ | $\begin{aligned} & 5060 \\ & 0 \end{aligned}$ | $\begin{aligned} & 73672 \\ & 6 \end{aligned}$ |
| $\begin{aligned} & 197 \\ & 9 \end{aligned}$ | 203300 | $\begin{aligned} & 2030 \\ & 0 \end{aligned}$ | $\begin{aligned} & 22360 \\ & 0 \end{aligned}$ | $\begin{aligned} & 39800 \\ & 0 \end{aligned}$ | $\begin{aligned} & 3980 \\ & 0 \end{aligned}$ | $\begin{aligned} & 43780 \\ & 0 \end{aligned}$ | 152323 | $\begin{aligned} & 50 \\ & 0 \end{aligned}$ | $\begin{aligned} & 15282 \\ & 3 \end{aligned}$ | 7000 |  | 7000 | 21932 |  | $\begin{aligned} & 2193 \\ & 2 \end{aligned}$ | $\begin{aligned} & 78255 \\ & 5 \end{aligned}$ | $\begin{aligned} & 6060 \\ & 0 \end{aligned}$ | $\begin{aligned} & 84315 \\ & 5 \end{aligned}$ |
| $\begin{aligned} & 198 \\ & 0 \end{aligned}$ | 218700 | 6000 | $\begin{aligned} & 22470 \\ & 0 \end{aligned}$ | $\begin{aligned} & 38610 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1560 \\ & 0 \end{aligned}$ | $\begin{aligned} & 40170 \\ & 0 \end{aligned}$ | 87931 |  | 87931 | 8300 |  | 8300 | 12280 |  | $\begin{aligned} & 1228 \\ & 0 \end{aligned}$ | $\begin{aligned} & 71331 \\ & 1 \end{aligned}$ | $\begin{aligned} & 2160 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & 73491 \\ & 1 \end{aligned}$ |
| $\begin{aligned} & 198 \\ & 1 \end{aligned}$ | 335100 | 2500 | $\begin{aligned} & 33760 \\ & 0 \end{aligned}$ | $\begin{aligned} & 27430 \\ & 0 \end{aligned}$ | $\begin{aligned} & 3980 \\ & 0 \end{aligned}$ | $\begin{aligned} & 31410 \\ & 0 \end{aligned}$ | 64172 | $\begin{aligned} & 32 \\ & 16 \end{aligned}$ | 67388 | 18700 |  | 18700 | 16688 |  | $\begin{aligned} & 1668 \\ & 8 \end{aligned}$ | $\begin{aligned} & 70896 \\ & 0 \end{aligned}$ | $\begin{aligned} & 4551 \\ & 6 \end{aligned}$ | $\begin{aligned} & 75447 \\ & 6 \end{aligned}$ |


| $\begin{gathered} \text { YEA } \\ R \end{gathered}$ | Subare $\text { A } 6$ | Subarea 7 and Divisions 8.abde |  |  | Subareas 3 AND 4 |  |  | SUbAREAS 125 AND 14 |  |  | Divisions 8.c AND 9.A |  | Total |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & 198 \\ & 2 \end{aligned}$ | 340400 | 4100 | $\begin{aligned} & 34450 \\ & 0 \end{aligned}$ | $\begin{aligned} & 25780 \\ & 0 \end{aligned}$ | $\begin{aligned} & 2080 \\ & 0 \end{aligned}$ | $\begin{aligned} & 27860 \\ & 0 \end{aligned}$ | 35033 | $\begin{aligned} & 45 \\ & 0 \end{aligned}$ | 35483 | 37600 | 37600 | 21076 | $\begin{aligned} & 2107 \\ & 6 \end{aligned}$ | $\begin{aligned} & 69190 \\ & 9 \end{aligned}$ | $\begin{aligned} & 2535 \\ & 0 \end{aligned}$ | $\begin{aligned} & 71725 \\ & 9 \end{aligned}$ |
| $\begin{aligned} & 198 \\ & 3 \end{aligned}$ | 320500 | 2300 | $\begin{aligned} & 32280 \\ & 0 \end{aligned}$ | $\begin{aligned} & 23500 \\ & 0 \end{aligned}$ | 9000 | $\begin{aligned} & 24400 \\ & 0 \end{aligned}$ | 40889 | 96 | 40985 | 49000 | 49000 | 14853 | $\begin{aligned} & 1485 \\ & 3 \end{aligned}$ | $\begin{aligned} & 66024 \\ & 2 \end{aligned}$ | $\begin{aligned} & 1139 \\ & 6 \end{aligned}$ | $\begin{aligned} & 67163 \\ & 8 \end{aligned}$ |
| $\begin{aligned} & 198 \\ & 4 \end{aligned}$ | 306100 | 1600 | $\begin{aligned} & 30770 \\ & 0 \end{aligned}$ | $\begin{aligned} & 16140 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1050 \\ & 0 \end{aligned}$ | $\begin{aligned} & 17190 \\ & 0 \end{aligned}$ | 43696 | $\begin{aligned} & 20 \\ & 2 \end{aligned}$ | 43898 | 98222 | 98222 | 20208 | $\begin{aligned} & 2020 \\ & 8 \end{aligned}$ | $\begin{aligned} & 62962 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1230 \\ & 2 \end{aligned}$ | $\begin{aligned} & 64192 \\ & 8 \end{aligned}$ |
| $\begin{aligned} & 198 \\ & 5 \end{aligned}$ | 388140 | 2735 | $\begin{aligned} & 39087 \\ & 5 \end{aligned}$ | 75043 | 1800 | 76843 | 46790 | $\begin{aligned} & 36 \\ & 56 \end{aligned}$ | 50446 | 78000 | 78000 | 18111 | $\begin{aligned} & 1811 \\ & 1 \end{aligned}$ | $\begin{aligned} & 60608 \\ & 4 \end{aligned}$ | 8191 | $\begin{aligned} & 61427 \\ & 5 \end{aligned}$ |
| $\begin{aligned} & 198 \\ & 6 \end{aligned}$ | 104100 |  | $\begin{aligned} & 10410 \\ & 0 \end{aligned}$ | $\begin{aligned} & 12849 \\ & 9 \end{aligned}$ |  | $\begin{aligned} & 12849 \\ & 9 \end{aligned}$ | 236309 | $\begin{aligned} & 74 \\ & 31 \end{aligned}$ | $\begin{aligned} & 24374 \\ & 0 \end{aligned}$ | 101000 | $\begin{aligned} & 10100 \\ & 0 \end{aligned}$ | 24789 | $\begin{aligned} & 2478 \\ & 9 \end{aligned}$ | $\begin{aligned} & 59469 \\ & 7 \end{aligned}$ | 7431 | $\begin{aligned} & 60212 \\ & 8 \end{aligned}$ |

Table 8.4.1.1. NE Atlantic Mackerel. ICES estimated catches by area ( $\mathbf{t}$ ). Discards not estimated prior to 1978 (data submitted by Working Group members). Continued.

| YeA | Subarea 6 |  |  | Subarea 7 and DIVISIons 8.AbDE |  |  | Subareas 3 AND 4 |  |  | Subareas 125 AND 14 |  |  | Divisions 8.c AND 9.A |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Dis <br> c | Catch | Ldg | Dis <br> c | Catc <br> h | Ldg | Disc | Catch |
| $\begin{aligned} & 198 \\ & 7 \end{aligned}$ | $\begin{aligned} & 18370 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & 18370 \\ & 0 \end{aligned}$ | $\begin{aligned} & 10030 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & 10030 \\ & 0 \end{aligned}$ | $\begin{aligned} & 29082 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1078 \\ & 9 \end{aligned}$ | $\begin{aligned} & 30161 \\ & 8 \end{aligned}$ | 47000 |  | 47000 | $\begin{aligned} & 2218 \\ & 7 \end{aligned}$ |  | $\begin{aligned} & 2218 \\ & 7 \end{aligned}$ | $\begin{aligned} & 64401 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1078 \\ & 9 \end{aligned}$ | $\begin{aligned} & 65480 \\ & 5 \end{aligned}$ |
| $\begin{aligned} & 198 \\ & 8 \end{aligned}$ | $\begin{aligned} & 11560 \\ & 0 \end{aligned}$ | 3100 | $\begin{aligned} & 11870 \\ & 0 \end{aligned}$ | 75600 | 2700 | 78300 | $\begin{aligned} & 30855 \\ & 0 \end{aligned}$ | $\begin{aligned} & 2976 \\ & 6 \end{aligned}$ | $\begin{aligned} & 33831 \\ & 6 \end{aligned}$ | $\begin{aligned} & 12040 \\ & 4 \end{aligned}$ |  | $\begin{aligned} & 12040 \\ & 4 \end{aligned}$ | $\begin{aligned} & 2477 \\ & 2 \end{aligned}$ |  | $\begin{aligned} & 2477 \\ & 2 \end{aligned}$ | $\begin{aligned} & 64492 \\ & 6 \end{aligned}$ | $\begin{aligned} & 3556 \\ & 6 \end{aligned}$ | $\begin{aligned} & 68049 \\ & 2 \end{aligned}$ |
| $\begin{aligned} & 198 \\ & 9 \end{aligned}$ | $\begin{aligned} & 12130 \\ & 0 \end{aligned}$ | 2600 | $\begin{aligned} & 12390 \\ & 0 \end{aligned}$ | 72900 | 2300 | 75200 | $\begin{aligned} & 27941 \\ & 0 \end{aligned}$ | 2190 | $\begin{aligned} & 28160 \\ & 0 \end{aligned}$ | 90488 |  | 90488 | $\begin{aligned} & 1832 \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 1832 \\ & 1 \end{aligned}$ | $\begin{aligned} & 58241 \\ & 9 \end{aligned}$ | 7090 | $\begin{aligned} & 58950 \\ & 9 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 0 \end{aligned}$ | $\begin{aligned} & 11480 \\ & 0 \end{aligned}$ | 5800 | $\begin{aligned} & 12060 \\ & 0 \end{aligned}$ | 56300 | 5500 | 61800 | $\begin{aligned} & 30080 \\ & 0 \end{aligned}$ | 4300 | $\begin{aligned} & 30510 \\ & 0 \end{aligned}$ | $\begin{aligned} & 11870 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & 11870 \\ & 0 \end{aligned}$ | $\begin{aligned} & 2131 \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 2131 \\ & 1 \end{aligned}$ | $\begin{aligned} & 61191 \\ & 1 \end{aligned}$ | $\begin{aligned} & 1560 \\ & 0 \end{aligned}$ | $\begin{aligned} & 62751 \\ & 1 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 1 \end{aligned}$ | $\begin{aligned} & 10950 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1070 \\ & 0 \end{aligned}$ | $\begin{aligned} & 12020 \\ & 0 \end{aligned}$ | 50500 | $\begin{aligned} & 1280 \\ & 0 \end{aligned}$ | 63300 | $\begin{aligned} & 35870 \\ & 0 \end{aligned}$ | 7200 | $\begin{aligned} & 36590 \\ & 0 \end{aligned}$ | 97800 |  | 97800 | $\begin{aligned} & 2068 \\ & 3 \end{aligned}$ |  | $\begin{aligned} & 2068 \\ & 3 \end{aligned}$ | $\begin{aligned} & 63718 \\ & 3 \end{aligned}$ | $\begin{aligned} & 3070 \\ & 0 \end{aligned}$ | $\begin{aligned} & 66788 \\ & 3 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 2 \end{aligned}$ | $\begin{aligned} & 14190 \\ & 6 \end{aligned}$ | 9620 | $\begin{aligned} & 15152 \\ & 6 \end{aligned}$ | 72153 | $\begin{aligned} & 1240 \\ & 0 \end{aligned}$ | 84553 | $\begin{aligned} & 36418 \\ & 4 \end{aligned}$ | 2980 | $\begin{aligned} & 36716 \\ & 4 \end{aligned}$ | $\begin{aligned} & 13906 \\ & 2 \end{aligned}$ |  | $\begin{aligned} & 13906 \\ & 2 \end{aligned}$ | $\begin{aligned} & 1804 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 1804 \\ & 6 \end{aligned}$ | $\begin{aligned} & 73535 \\ & 1 \end{aligned}$ | $\begin{aligned} & 2500 \\ & 0 \end{aligned}$ | $\begin{aligned} & 76035 \\ & 1 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 3 \end{aligned}$ | $\begin{aligned} & 13349 \\ & 7 \end{aligned}$ | 2670 | $\begin{aligned} & 13616 \\ & 7 \end{aligned}$ | 99828 | $\begin{aligned} & 1279 \\ & 0 \end{aligned}$ | $\begin{aligned} & 11261 \\ & 8 \end{aligned}$ | $\begin{aligned} & 38783 \\ & 8 \end{aligned}$ | 2720 | $\begin{aligned} & 39055 \\ & 8 \end{aligned}$ | $\begin{aligned} & 16597 \\ & 3 \end{aligned}$ |  | $\begin{aligned} & 16597 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1972 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & 1972 \\ & 0 \end{aligned}$ | $\begin{aligned} & 80685 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1818 \\ & 0 \end{aligned}$ | $\begin{aligned} & 82503 \\ & 6 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 4 \end{aligned}$ | $\begin{aligned} & 13433 \\ & 8 \end{aligned}$ | 1390 | $\begin{aligned} & 13572 \\ & 8 \end{aligned}$ | $\begin{aligned} & 11308 \\ & 8 \end{aligned}$ | 2830 | $\begin{aligned} & 11591 \\ & 8 \end{aligned}$ | $\begin{aligned} & 47124 \\ & 7 \end{aligned}$ | 1150 | $\begin{aligned} & 47239 \\ & 7 \end{aligned}$ | 72309 |  | 72309 | $\begin{aligned} & 2504 \\ & 3 \end{aligned}$ |  | $\begin{aligned} & 2504 \\ & 3 \end{aligned}$ | $\begin{aligned} & 81602 \\ & 5 \end{aligned}$ | 5370 | $\begin{aligned} & 82139 \\ & 5 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 5 \end{aligned}$ | $\begin{aligned} & 14562 \\ & 6 \end{aligned}$ | 74 | $\begin{aligned} & 14570 \\ & 0 \end{aligned}$ | $\begin{aligned} & 11788 \\ & 3 \end{aligned}$ | 6917 | $\begin{aligned} & 12480 \\ & 0 \end{aligned}$ | $\begin{aligned} & 32147 \\ & 4 \end{aligned}$ | 730 | $\begin{aligned} & 32220 \\ & 4 \end{aligned}$ | $\begin{aligned} & 13549 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 13549 \\ & 6 \end{aligned}$ | $\begin{aligned} & 2760 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & 2760 \\ & 0 \end{aligned}$ | $\begin{aligned} & 74807 \\ & 9 \end{aligned}$ | 7721 | $\begin{aligned} & 75580 \\ & 0 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 6 \end{aligned}$ | $\begin{aligned} & 12989 \\ & 5 \end{aligned}$ | 255 | $\begin{aligned} & 13015 \\ & 0 \end{aligned}$ | 73351 | 9773 | 83124 | $\begin{aligned} & 21145 \\ & 1 \end{aligned}$ | 1387 | $\begin{aligned} & 21283 \\ & 8 \end{aligned}$ | $\begin{aligned} & 10337 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 10337 \\ & 6 \end{aligned}$ | $\begin{aligned} & 3412 \\ & 3 \end{aligned}$ |  | $\begin{aligned} & 3412 \\ & 3 \end{aligned}$ | $\begin{aligned} & 55219 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1141 \\ & 5 \end{aligned}$ | $\begin{aligned} & 56361 \\ & 1 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 7 \end{aligned}$ | 65044 | 2240 | 67284 | $\begin{aligned} & 11471 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1381 \\ & 7 \end{aligned}$ | $\begin{aligned} & 12853 \\ & 6 \end{aligned}$ | $\begin{aligned} & 22668 \\ & 0 \end{aligned}$ | 2807 | $\begin{aligned} & 22948 \\ & 7 \\ & \hline \end{aligned}$ | $\begin{aligned} & 10359 \\ & 8 \end{aligned}$ |  | $\begin{aligned} & 10359 \\ & 8 \end{aligned}$ | $\begin{aligned} & 4070 \\ & 8 \end{aligned}$ |  | $\begin{aligned} & 4070 \\ & 8 \end{aligned}$ | $\begin{aligned} & 55074 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1886 \\ & 4 \end{aligned}$ | $\begin{aligned} & 56961 \\ & 3 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 8 \end{aligned}$ | $\begin{aligned} & 11014 \\ & 1 \end{aligned}$ | 71 | $\begin{aligned} & 11021 \\ & 2 \end{aligned}$ | $\begin{aligned} & 10518 \\ & 1 \end{aligned}$ | 3206 | $\begin{aligned} & 10838 \\ & 7 \end{aligned}$ | $\begin{aligned} & 26494 \\ & 7 \end{aligned}$ | 4735 | $\begin{aligned} & 26968 \\ & 2 \end{aligned}$ | $\begin{aligned} & 13421 \\ & 9 \end{aligned}$ |  | $\begin{aligned} & 13421 \\ & 9 \end{aligned}$ | $\begin{aligned} & 4416 \\ & 4 \end{aligned}$ |  | $\begin{aligned} & 4416 \\ & 4 \end{aligned}$ | $\begin{aligned} & 65865 \\ & 2 \end{aligned}$ | 8012 | $\begin{aligned} & 66666 \\ & 4 \end{aligned}$ |
| $\begin{aligned} & 199 \\ & 9 \end{aligned}$ | $\begin{aligned} & 11636 \\ & 2 \end{aligned}$ |  | $\begin{aligned} & 11636 \\ & 2 \end{aligned}$ | 94290 |  | 94290 | $\begin{aligned} & 31301 \\ & 4 \end{aligned}$ |  | $\begin{aligned} & 31301 \\ & 4 \end{aligned}$ | 72848 |  | 72848 | $\begin{aligned} & 4379 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 4379 \\ & 6 \end{aligned}$ | $\begin{aligned} & 64031 \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 64031 \\ & 1 \end{aligned}$ |


| $\begin{aligned} & \text { YEA } \\ & R \end{aligned}$ | Subarea 6 |  |  | Subarea 7 and Divisions 8.abde |  |  | Subareas 3 AND 4 |  |  | SUbAREAS 125 AND 14 |  |  | Divisions 8.c AND 9.A |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & 200 \\ & 0 \end{aligned}$ | $\begin{aligned} & 18759 \\ & 5 \end{aligned}$ | 1 | $\begin{aligned} & 18759 \\ & 5 \end{aligned}$ | $\begin{aligned} & 11556 \\ & 6 \end{aligned}$ | 1918 | $\begin{aligned} & 11748 \\ & 4 \end{aligned}$ | $\begin{aligned} & 28556 \\ & 7 \end{aligned}$ | 165 | $\begin{aligned} & 30489 \\ & 8 \end{aligned}$ | 92557 |  | 92557 | $\begin{aligned} & 3607 \\ & 4 \end{aligned}$ |  | $\begin{aligned} & 3607 \\ & 4 \end{aligned}$ | $\begin{aligned} & 73652 \\ & 4 \end{aligned}$ | 2084 | $\begin{aligned} & 73860 \\ & 8 \end{aligned}$ |
| $\begin{aligned} & 200 \\ & 1 \end{aligned}$ | $\begin{aligned} & 14314 \\ & 2 \end{aligned}$ | 83 | $\begin{aligned} & 14314 \\ & 2 \end{aligned}$ | $\begin{aligned} & 14289 \\ & 0 \end{aligned}$ | 1081 | $\begin{aligned} & 14397 \\ & 1 \end{aligned}$ | $\begin{aligned} & 32720 \\ & 0 \end{aligned}$ | 24 | $\begin{aligned} & 33997 \\ & 1 \end{aligned}$ | 67097 |  | 67097 | $\begin{aligned} & 4319 \\ & 8 \end{aligned}$ |  | $\begin{aligned} & 4319 \\ & 8 \end{aligned}$ | $\begin{aligned} & 73627 \\ & 4 \end{aligned}$ | 1188 | $\begin{aligned} & 73746 \\ & 2 \end{aligned}$ |
| $\begin{aligned} & 200 \\ & 2 \end{aligned}$ | $\begin{aligned} & 13684 \\ & 7 \end{aligned}$ | $\begin{aligned} & 1293 \\ & 1 \end{aligned}$ | $\begin{aligned} & 14977 \\ & 8 \end{aligned}$ | $\begin{aligned} & 10248 \\ & 4 \end{aligned}$ | 2260 | $\begin{aligned} & 10474 \\ & 4 \end{aligned}$ | $\begin{aligned} & 37570 \\ & 8 \end{aligned}$ | 8583 | $\begin{aligned} & 39487 \\ & 8 \end{aligned}$ | 73929 |  | 73929 | $\begin{aligned} & 4957 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 4957 \\ & 6 \end{aligned}$ | $\begin{aligned} & 74913 \\ & 1 \end{aligned}$ | $\begin{aligned} & 2377 \\ & 4 \end{aligned}$ | $\begin{aligned} & 77290 \\ & 5 \end{aligned}$ |
| $\begin{aligned} & 200 \\ & 3 \end{aligned}$ | $\begin{aligned} & 13569 \\ & 0 \end{aligned}$ | 1399 | $\begin{aligned} & 13708 \\ & 9 \end{aligned}$ | 90356 | 5712 | 96068 | $\begin{aligned} & 35410 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1178 \\ & 5 \end{aligned}$ | $\begin{aligned} & 36589 \\ & 4 \end{aligned}$ | 53883 |  | 53883 | $\begin{aligned} & 2582 \\ & 3 \end{aligned}$ | 531 | $\begin{aligned} & 2635 \\ & 4 \end{aligned}$ | $\begin{aligned} & 65983 \\ & 1 \end{aligned}$ | $\begin{aligned} & 1942 \\ & 7 \end{aligned}$ | $\begin{aligned} & 67928 \\ & 8 \end{aligned}$ |
| $\begin{aligned} & 200 \\ & 4 \end{aligned}$ | $\begin{aligned} & 13403 \\ & 3 \end{aligned}$ | 1705 | $\begin{aligned} & 13473 \\ & 8 \end{aligned}$ | $\begin{aligned} & 10370 \\ & 3 \end{aligned}$ | 5991 | $\begin{aligned} & 10969 \\ & 4 \end{aligned}$ | $\begin{aligned} & 30604 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1132 \\ & 9 \end{aligned}$ | $\begin{aligned} & 31736 \\ & 9 \end{aligned}$ | 62913 | 9 | 62922 | $\begin{aligned} & 3484 \\ & 0 \end{aligned}$ | 928 | $\begin{aligned} & 3576 \\ & 9 \end{aligned}$ | $\begin{aligned} & 64052 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1996 \\ & 2 \end{aligned}$ | $\begin{aligned} & 66049 \\ & 1 \end{aligned}$ |
| $\begin{aligned} & 200 \\ & 5 \end{aligned}$ | 79960 | 8201 | 88162 | 90278 | $\begin{aligned} & 1215 \\ & 8 \end{aligned}$ | $\begin{aligned} & 10243 \\ & 6 \end{aligned}$ | $\begin{aligned} & 24974 \\ & 1 \end{aligned}$ | 4633 | $\begin{aligned} & 25437 \\ & 4 \end{aligned}$ | 54129 |  | 54129 | $\begin{aligned} & 4961 \\ & 8 \end{aligned}$ | 796 | $\begin{aligned} & 5041 \\ & 4 \end{aligned}$ | $\begin{aligned} & 52372 \\ & 6 \end{aligned}$ | $\begin{aligned} & 2578 \\ & 8 \end{aligned}$ | $\begin{aligned} & 54951 \\ & 4 \end{aligned}$ |

Table 8.4.1.1. NE Atlantic Mackerel. ICES estimated catches by area ( $\mathbf{t}$ ). Discards not estimated prior to 1978 (data submitted by Working Group members). Continued.

| Year | Subarea 6 |  |  | Subarea 7 and <br> DIVISIONS 8.AbDE |  |  | Subareas 3 AND 4 |  |  | Subareas 125 AND 14 |  |  | Divisions 8.c AND 9.A |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch | Ldg | Disc | Catch |
| 2006 | 88077 | 6081 | 94158 | 66209 | 8642 | 74851 | 200929 | 8263 | 209192 | 46716 |  | 46716 | 52751 | 3607 | 56358 | 454587 | 26594 | 481181 |
| 2007 | 110788 | 2450 | 113238 | 71235 | 7727 | 78962 | 253013 | 4195 | 257208 | 72891 |  | 72891 | 62834 | 1072 | 63906 | 570762 | 15444 | 586206 |
| 2008 | 76358 | 21889 | 98247 | 73954 | 5462 | 79416 | 227252 | 8862 | 236113 | 148669 | 112 | 148781 | 59859 | 750 | 60609 | 586090 | 37075 | 623165 |
| 2009 | 135468 | 3927 | 139395 | 88287 | 2921 | 91208 | 226928 | 8120 | 235049 | 163604 |  | 163604 | 107747 | 966 | 108713 | 722035 | 15934 | 737969 |
| 2010 | 106732 | 2904 | 109636 | 104128 | 4614 | 108741 | 246818 | 883 | 247700 | 355725 | 5 | 355729 | 49068 | 4640 | 53708 | 862470 | 13045 | 875515 |
| 2011 | 160756 | 1836 | 162592 | 51098 | 5317 | 56415 | 301746 | 1906 | 303652 | 398132 | 28 | 398160 | 24036 | 1807 | 25843 | 935767 | 10894 | 946661 |
| 2012 | 121115 | 952 | 122067 | 65728 | 9701 | 75429 | 218400 | 1089 | 219489 | 449325 | 1 | 449326 | 24941 | 3431 | 28372 | 879510 | 15174 | 894684 |
| 2013 | 132062 | 273 | 132335 | 49871 | 1652 | 51523 | 260921 | 337 | 261258 | 465714 | 15 | 465729 | 19733 | 2455 | 22188 | 928433 | 4732 | 933165 |
| 2014 | 180068 | 340 | 180408 | 93709 | 1402 | 95111 | 383887 | 334 | 384221 | 684082 | 91 | 684173 | 46257 | 4284 | 50541 | 1388003 | 6451 | 1394454 |
| 2015 | 134728 | 30 | 134757 | 98563 | 3155 | 101718 | 295877 | 34 | 295911 | 632493 | 78 | 632571 | 36899 | 7133 | 44033 | 1198560 | 10431 | 1208990 |
| 2016 | 206326 | 200 | 206526 | 37300 | 1927 | 39227 | 248041 | 570 | 248611 | 563440 | 54 | 563494 | 32987 | 3220 | 36207 | 1088094 | 5971 | 1094066 |
| 2017 | 225959 | 151 | 226110 | 21128 | 1992 | 23119 | 269404 | 400 | 269804 | 603806 | 62 | 603869 | 32815 | 227 | 33042 | 1153112 | 2832 | 1155944 |

Table 8.4.2.1. NE Atlantic Mackerel. ICES estimated catch (t) in Subareas 1, 2, 5 and 14, 1984-2017 (Data submitted by Working Group members).

| Country | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 11787 | 7610 | 1653 | 3133 | 4265 | 6433 | 6800 | 1098 | 251 |
| Estonia |  |  |  |  |  |  |  |  | 216 |
| Faroe <br> Islands | 137 |  |  |  | 22 | 1247 | 3100 | 5793 | 3347 |
| France |  | 16 |  |  |  | 11 |  | 23 | 6 |
| Germany <br> Fed. Rep. |  |  | 99 |  | 380 |  |  |  |  |
| Germany <br> Dem. Rep. |  |  | 16 | 292 |  | 2409 |  |  |  |
| Iceland |  |  |  |  |  |  |  |  |  |
| Ireland |  |  |  |  |  |  |  |  |  |
| Latvia |  |  |  |  |  |  |  |  | 100 |
| Lithuania |  |  |  |  |  |  |  |  |  |
| Netherlands |  |  |  |  |  |  |  |  |  |
| Norway | 82005 | 61065 | 85400 | 25000 | 86400 | 68300 | 77200 | 76760 | 91900 |
| Poland |  |  |  |  |  |  |  |  |  |
| Sweden |  |  |  |  |  |  |  |  |  |
| United <br> Kingdom |  |  | 2131 | 157 | 1413 |  | 400 | 514 | 802 |
| USSR/Russia | 4293 | 9405 | 11813 | 18604 | 27924 | 12088 | 28900 | 13361 | 42440 |
| Misreported <br> (Area 4.a) |  |  |  |  |  |  |  |  |  |
| Misreported <br> (Area 6.a) |  |  |  |  |  |  |  |  |  |
| Misreported (Unknown) |  |  |  |  |  |  |  |  |  |
| Unallocated |  |  |  |  |  |  |  |  |  |
| Discards |  |  |  |  |  |  |  |  |  |
| Total | 98222 | 78096 | 101112 | 47186 | 120404 | 90488 | 118700 | 97819 | 139062 |

Table 8.4.2.1. NE Atlantic Mackerel. ICES estimated catch ( $t$ ) in Areas 1, 2,5 and 14, 1984-2017. Continued.

| Country | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  | 4746 | 3198 | 37 | 2090 | 106 | 1375 |
| Estonia |  | 3302 | 1925 | 3741 | 4422 | 7356 | 3595 | 2673 |
| Faroe <br> Islands | 1167 | 6258 | 9032 | 2965 | 5777 | 2716 | 3011 | 5546 |
| France | 6 | 5 | 5 |  | 270 |  |  |  |
| Germany |  |  |  |  |  |  |  |  |
| Greenland |  |  |  | 1 |  |  |  |  |
| Iceland |  |  |  | 92 | 925 | 357 |  |  |
| Ireland |  |  |  |  |  |  | 100 |  |
| Latvia | 4700 | 1508 | 389 | 233 |  |  |  |  |
| Lithuania |  |  |  |  |  |  |  | 2085 |
| Netherlands |  |  |  | 561 |  |  | 661 |  |
| Norway | 100500 | 141114 | 93315 | 47992 | 41000 | 54477 | 53821 | 31778 |
| Poland |  |  |  |  | 22 |  |  |  |
| Sweden |  |  |  |  |  |  |  |  |
| United Kingdom |  | 1706 | 194 | 48 | 938 | 199 | 662 |  |
| Russia | 49600 | 28041 | 44537 | 44545 | 50207 | 67201 | 51003 | 491001 |
| Misreported (Area 4.a) |  | -109625 | -18647 |  |  | -177 | -40011 |  |
| Misreported (Area 6.a) |  |  |  |  |  |  | -100 |  |
| Misreported (Unknown) |  |  |  |  |  |  |  |  |
| Unallocated |  |  |  |  |  |  |  |  |
| Discards |  |  |  |  |  |  |  |  |
| Total | 165973 | 72309 | 135496 | 103376 | 103598 | 134219 | 72848 | 92557 |

Table 8.4.2.1. NE Atlantic Mackerel. ICES estimated catch (t) in Areas 1, 2, 5, and 14, 1984-2017. Continued.

| Country | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 7 | 1 |  |  |  |  |  |  |
| Estonia | 219 |  |  |  |  |  |  |  |
| Faroe Islands | 3272 | 4730 |  | 650 | 30 |  | 278 | 123 |
| France |  |  |  | 2 | 1 |  |  |  |
| Germany |  |  |  |  |  |  | 7 |  |
| Greenland |  |  |  |  |  |  |  |  |
| Iceland |  | 53 | 122 |  | 363 | 4222 | 36706 | 112286 |
| Ireland |  |  | 495 | 471 |  |  |  |  |
| Latvia |  |  |  |  |  |  |  |  |
| Lithuania |  |  |  |  |  |  |  |  |
| Netherlands |  | 569 | 44 | 34 | 2393 |  | 10 | 72 |
| Norway | 21971 | 22670 | 125481 | 10295 | 13244 | 8914 | 493 | 3474 |
| Poland |  |  |  |  |  |  |  |  |
| Sweden | 8 |  |  |  |  |  |  |  |
| United <br> Kingdom | 54 | 665 | 692 | 2493 |  |  |  | 4 |
| Russia | 41566 | 45811 | 40026 | 49489 | 40491 | 33580 | 35408 | 32728 |
| Misreported (Area 4.a) |  |  |  |  |  |  |  |  |
| Misreported (Area 6.a) |  |  |  |  |  |  |  |  |
| Misreported (Unknown) |  | -570 |  | -553 |  |  |  |  |
| Unallocated |  |  | -44 | 32 | -2393 |  | -10 | -18 |
| Discards |  |  |  | 9 |  |  |  | 112 |
| Total | 67097 | 73929 | 53883 | 62922 | 54129 | 46716 | 72891 | 148781 |

Table 8.4.2.1. NE Atlantic Mackerel. ICES estimated catch (t) in Areas 1, 2, 5, and 14, 1984-2017. Continued.

| Country | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  | 4845 | 269 |  | 391 | 2345 | 4321 | 1 | 2 |
| Estonia |  |  |  |  | 13671 |  | 0 |  |  |
| Faroe Islands | 2992 | 66312 | 121499 | 107198 | 142976 | 103896 | 76889 | 61901 | 66194 |
| France |  |  | 2 |  | 197 | 8 | 36 |  |  |
| Germany |  |  |  | 107 | 74 |  | 2963 | 3499 | 4064 |
| Greenland |  |  | 621 | 74021 | 541481 | 875811 | 30351 | 36142 | 46388 |
| Iceland | 116160 | 121008 | 159263 | 149282 | 151103 | 172960 | 169333 | 170374 | 167366 |
| Ireland |  |  | 90 |  |  | 1725 | 6 | 2 |  |
| Latvia |  |  |  |  |  |  |  |  |  |
| Lithuania |  |  |  |  |  | 1082 |  | 1931 |  |
| Netherlands |  | 90 | 178 | 5 | 1 | 5887 | 6996 | 8599 | 7671 |
| Norway | 3038 | 104858 | 43168 | 110741 | 33817 | 192322 | 204574 | 153228 | 167739 |
| Poland |  |  |  |  |  |  |  |  |  |
| Sweden |  |  |  | 4 | 825 | 3310 | 740 | 730 | 1720 |
| United Kingdom |  |  |  |  | 2 | 5534 | 7851 | 5240 | 4601 |
| Russia | 414141 | 58613 | 73601 | 74587 | 80812 | 116433 | 128433 | 121614 | 138061 |
| Misreported (Area 4.a) |  |  |  |  |  |  |  |  |  |
| Misreported (Area 6.a) |  |  |  |  |  |  |  |  |  |
| Misreported (Unknown) |  |  |  |  |  |  |  |  |  |
| Unallocated |  |  |  |  |  |  |  |  |  |
| Discards |  | 5 | 28 | 1 | 151 | 911 | 78 | 54 | 62 |
| Total | 163604 | 355729 | 398160 | 449326 | 465729 | 684173 | 632571 | 563315 | 603869 |

Table 8.4.2.2. NE Atlantic Mackerel. ICES estimated catch ( $\mathbf{t}$ ) in the North Sea, Skagerrak and Kattegat (Subarea 4 and Division 3.a), 1988-2017 (Data submitted by Working Group members).

| CountRY | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | 20 | 37 |  | 125 | 102 | 191 | 351 | 106 |
| Denmark | 32588 | 26831 | 29000 | 38834 | 41719 | 42502 | 47852 | 30891 |
| Estonia |  |  |  |  | 400 |  |  |  |
| Faroe <br> Islands |  | 2685 | 5900 | 5338 |  | 11408 | 11027 | 17883 |
| France | 1806 | 2200 | 1600 | 2362 | 956 | 1480 | 1570 | 1599 |
| Germany <br> Fed. Rep. | 177 | 6312 | 3500 | 4173 | 4610 | 4940 | 1497 | 712 |
| Iceland |  |  |  |  |  |  |  |  |
| Ireland |  |  |  |  |  |  |  |  |

Table 8.4.2.2. NE Atlantic Mackerel. ICES estimated catch ( t ) in the North Sea, Skagerrak and Kattegat (Sub-area 4 and Division 3.a), 1988-2017. Continued.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 62 | 114 | 125 | 177 | 146 | 97 | 22 |
| Denmark | 24057 | 21934 | 25326 | 29353 | 27720 | 21680 | 343751 |
| Estonia |  |  |  |  |  |  |  |
| Faroe Islands | 13886 | 32882 | 4832 | 4370 | 10614 | 18751 | 12548 |
| France | 1316 | 1532 | 1908 | 2056 | 1588 | 1981 | 2152 |
| Germany | 542 | 213 | 423 | 473 | 78 | 4514 | 3902 |
| Iceland |  |  |  | 357 |  |  |  |
| Ireland | 5280 | 280 | 145 | 11293 | 9956 | 10284 | 20715 |
| Latvia |  |  |  |  |  |  |  |
| Lithuania |  |  |  |  |  |  |  |
| Netherlands | 1996 | 951 | 1373 | 2819 | 2262 | 2441 | 11044 |
| Norway | 88444 | 96300 | 103700 | 106917 | 142320 | 158401 | 161621 |
| Poland |  |  |  |  |  |  |  |
| Romania |  |  |  |  |  |  |  |
| Sweden | 5307 | 4714 | 5146 | 5233 | 49941 | 5090 | 52321 |
| United <br> Kingdom | 18545 | 19204 | 19755 | 32396 | 58282 | 52988 | 61781 |
| Russia |  | 3525 | 635 | 345 | 1672 | 1 |  |
| Misreported (Area 2.a) |  |  |  | 40000 |  |  |  |
| Misreported <br> (Area 6.a) | 51781 | 73523 | 98432 | 59882 | 8591 | 39024 | 49918 |
| Misreported (Unknown) |  |  |  |  |  |  |  |
| Unallocated | 236 | 1102 | 3147 | 17344 | 34761 | 24873 | 22985 |
| Discards | 1387 | 2807 | 4753 |  | 1912 | 24 | 8583 |
| Total | 212839 | 229487 | 269700 | 313015 | 304896 | 339970 | 394878 |

Table 8.4.2.2. NE Atlantic Mackerel. ICES estimated catch ( $\mathbf{t}$ ) in the North Sea, Skagerrak and Kattegat (Subarea 4 and Division 3.a), 1988-2017. Continued.

| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 2 | 4 | 1 | 3 | 1 | 2 | 3 |
| Denmark | 275081 | 25665 | 232121 | 242191 | 252171 | 26716 | 23491 |
| Estonia |  |  |  |  |  |  |  |
| Faroe Islands | 11754 | 11705 | 9739 | 12008 | 11818 | 7627 | 6648 |
| France | 1467 | 1538 | 1004 | 285 | 7549 | 490 | 1493 |
| Germany | 4859 | 4515 | 4442 | 2389 | 5383 | 4668 | 5158 |
| Iceland |  |  |  |  |  |  |  |
| Ireland | 17145 | 18901 | 15605 | 4125 | 13337 | 11628 | 12901 |
| Latvia |  |  |  |  |  |  |  |
| Lithuania |  |  |  |  |  |  |  |
| Netherlands | 6784 | 6366 | 3915 | 4093 | 5973 | 1980 | 2039 |
| Norway | 150858 | 147068 | 106434 | 113079 | 131191 | 114102 | 118070 |
| Poland |  |  | 109 |  |  |  |  |
| Romania |  |  |  |  |  |  |  |
| Sweden | 4450 | 4437 | 3204 | 3209 | 38581 | 36641 | 73031 |
| United Kingdom | 67083 | 62932 | 37118 | 28628 | 46264 | 37055 | 47863 |
| Russia |  |  | 4 |  |  |  |  |
| Misreported (Area 2.a) |  |  |  |  |  |  |  |
| Misreported (Area 6.a) | 62928 | 23692 | 37911 | 8719 |  | 17280 | 1959 |
| Misreported (Unknown) |  |  |  |  |  |  |  |
| Unallocated | -730 | -783 | 7043 | 171 | 2421 | 2039 | -629 |
| Discards | 11785 | 11329 | 4633 | 8263 | 4195 | 8862 | 8120 |
| Total | 365894 | 317369 | 254374 | 209192 | 257208 | 236111 | 235049 |

Table 8.4.2.2. NE Atlantic Mackerel. ICES estimated catch ( $t$ ) in the North Sea, Skagerrak and Kattegat (Subarea 4 and Division 3.a), 1988-2017. Continued.

| Country | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 27 | 21 | 39 | 62 | 56 | 38 | 99 | 107 |
| Denmark | 36552 | 32800 | 36492 | 31924 | 21340 | 35809 | 21696 | 27457 |
| Estonia |  |  |  |  |  |  |  |  |
| Faroe Islands | 4639 | 543 | 432 | 25 | 42919 | 25672 | 18193 | 12915 |
| France | 686 | 1416 | 5736 | 1788 | 4912 | 7827 | 3448 | 5942 |
| Germany | 25621 | 52911 | 4560 | 5755 | 4979 | 6056 | 10172 | 11185 |
| Iceland |  |  |  |  |  |  |  |  |
| Ireland | 14639 | 15810 | 20422 | 13523 | 45167 | 34167 | 24437 | 35957 |
| Latvia |  |  |  |  |  |  |  |  |
| Lithuania |  |  |  |  | 8340 |  | 596 |  |
| Netherlands | 1300 | 9881 | 6018 | 4863 | 24536 | 17547 | 11434 | 17401 |
| Norway | 129064 | 162878 | 64181 | 130056 | 85409 | 36344 | 55089 | 51960 |
| Poland |  |  |  |  |  | 24 |  | 0.721 |
| Romania |  |  |  |  |  |  |  |  |
| Sweden | 34291 | 32481 | 4560 | 2081 | 1112 | 3190 | 2933 | 1981 |
| United Kingdom | 52563 | 69858 | 75959 | 70840 | 145119 | 129203 | 99945 | 104499 |
| Russia | 696 |  |  | 4 |  |  |  |  |
| Misreported (Area 2.a) |  |  |  |  |  |  |  |  |
| Misreported (Area 6.a) |  |  |  |  |  |  |  |  |
| Misreported (Unknown) |  |  |  |  |  |  |  |  |
| Unallocated | 660 |  |  |  |  |  |  |  |
| Discards | 883 | 1906 | 1089 | 337 | 334 | 34 | 559 | 400 |
| Total | 247700 | 303652 | 219489 | 261258 | 384221 | 295911 | 248611 | 269804 |

Table 8.4.2.3. NE Atlantic Mackerel. ICES estimated catch ( $\mathbf{t}$ ) in the Western area (Sub-areas 6 and 7 and Divisions 8.a,b,d,e), 1985-2017 (Data submitted by Working Group members).

| Country | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  |  |  |  |  |
| Denmark | 400 | 300 | 100 |  | 1000 |  | 1573 | 194 |
| Estonia |  |  |  |  |  |  |  |  |
| Faroe <br> Islands | 9900 | 1400 | 7100 | 2600 | 1100 | 1000 |  |  |
| France | 7400 | 11200 | 11100 | 8900 | 12700 | 17400 | 4095 |  |
| Germany | 11800 | 7700 | 13300 | 15900 | 16200 | 18100 | 10364 | 9109 |
| Guernsey |  |  |  |  |  |  |  |  |
| Ireland | 91400 | 74500 | 89500 | 85800 | 61100 | 61500 | 17138 | 21952 |
| Isle of Man |  |  |  |  |  |  |  |  |
| Jersey |  |  |  |  |  |  |  |  |
| Lithuania |  |  |  |  |  |  |  |  |
| Netherlands | 37000 | 58900 | 31700 | 26100 | 24000 | 24500 | 64827 | 76313 |
| Norway | 24300 | 21000 | 21600 | 17300 | 700 |  | 29156 | 32365 |
| Poland |  |  |  |  |  |  |  |  |
| Spain |  |  |  | 1500 | 1400 | 400 | 4020 | 2764 |
| United <br> Kingdom | 205900 | 156300 | 200700 | 208400 | 149100 | 162700 | 162588 | 196890 |
| Misreported <br> (Area 4.a) |  | -148000 | -117000 | -180000 | -92000 | -126000 | -130000 | -127000 |
| Misreported <br> (Unknown) |  |  |  |  |  |  |  |  |
| Unallocated | 75100 | 49299 | 26000 | 4700 | 18900 | 11500 | -3802 | 1472 |
| Discards | 4500 |  |  | 5800 | 4900 | 11300 | 23550 | 22020 |
| Total | 467700 | 232599 | 284100 | 197000 | 199100 | 182400 | 183509 | 236079 |

Table 8.4.2.3. NE Atlantic Mackerel. ICES estimated catch (t) in the Western area (Sub-areas 6 and 7 and Divisions 8.a,b,d,e), 1985-2017 (Data submitted by Working Group members).

| CountRY | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium |  |  |  |  |  |  |  |  |
| Denmark |  | 2239 | 1143 | 1271 |  |  | 552 | 82 |
| Estonia |  |  | 361 |  |  |  |  |  |
| Faroe Islands |  | 4283 | 4284 |  | 24481 | 3681 | 4239 | 4863 |
| France | 2350 | 9998 | 10178 | 14347 | 19114 | 15927 | 14311 | 17857 |
| Germany | 8296 | 25011 | 23703 | 15685 | 15161 | 20989 | 19476 | 22901 |
| Guernsey |  |  |  |  |  |  |  |  |
| Ireland | 23776 | 79996 | 72927 | 49033 | 52849 | 66505 | 48282 | 61277 |
| Isle of Man |  |  |  |  |  |  |  |  |
| Jersey |  |  |  |  |  |  |  |  |
| Lithuania |  |  |  |  |  |  |  |  |
| Netherlands | 81773 | 40698 | 34514 | 34203 | 22749 | 28790 | 25141 | 30123 |
| Norway | 44600 | 2552 |  |  |  |  |  |  |
| Poland | 600 |  |  |  |  |  |  |  |
| Spain | 3162 | 4126 | 4509 | 2271 | 7842 | 3340 | 4120 | 4500 |
| United <br> Kingdom | 215265 | 208656 | 190344 | 127612 | 128836 | 165994 | 127094 | 126620 |
| Misreported <br> (Area 4.a) | -146697 | -134765 | -106987 | -51781 | -73523 | -98255 | -59982 | -3775 |
| Misreported <br> (Unknown) |  |  | 4632 | 28245 | 10603 | 4577 | 8351 | 21652 |
| Unallocated | 15660 | 4220 | 6991 | 10028 | 16057 | 3277 |  | 1920 |
| Discards | 248785 | 251646 | 270212 | 213272 | 196110 | 218599 | 204885 | 297932 |
| Total |  |  |  |  |  |  |  |  |

Table 8.4.2.3. NE Atlantic Mackerel. ICES estimated catch (t) in the Western area (Sub-areas 6 and 7 and Divisions 8.a,b,d,e), 1985-2017. Continued.

| CountRy | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium |  |  |  | 1 |  |  |  |  |
| Denmark | 835 |  | 113 |  |  |  | 6 | 10 |
| Estonia |  |  |  |  |  |  |  |  |
| Faroe <br> Islands | 2161 | 2490 | 2260 | 674 |  | 59 | 1333 | 3539 |
| France | 18975 | 19726 | 21213 | 18549 | 15182 | 14625 | 12434 | 14944 |
| Germany | 20793 | 22630 | 19200 | 18730 | 14598 | 14219 | 12831 | 10834 |
| Guernsey |  |  |  |  |  | 10 |  |  |
| Ireland | 60168 | 51457 | 49715 | 41730 | 30082 | 36539 | 35923 | 33132 |
| Isle of Man |  |  |  |  |  | 9 | 8 | 6 |
| Jersey |  |  |  |  |  |  |  | 95 |
| Lithuania |  |  |  |  |  |  |  |  |
| Netherlands | 33654 | 21831 | 23640 | 21132 | 18819 | 20064 | 18261 | 17920 |
| Norway |  |  |  |  |  |  |  | 1368 |
| Poland |  |  |  |  |  |  |  |  |
| Russia |  |  |  |  |  |  |  |  |
| Spain | 4063 | 3483 |  |  |  |  |  |  |
| United <br> Kingdom | 139589 | 131599 | 167246 | 149346 | 115586 | 67187 | 87424 | 768821 |
| Misreported <br> (Area 4.a) | -39024 | -43339 | -62928 | -23139 | -37911 | -8719 |  | -17280 |
| Misreported <br> (Unknown) |  |  |  |  |  |  |  |  |
| Unallocated | 37952 | 27558 | 5587 | 9714 | 13412 | 4783 | 10042 | -952 |
| Discards | 1164 | 15191 | 7111 | 7696 | 20359 | 14723 | 10177 | 27351 |
| Total | 280553 | 252620 | 233157 | 244432 | 190597 | 169009 | 192201 | 177662 |

Table 8.4.2.3. NE Atlantic Mackerel. ICES estimated catch ( $t$ ) in the Western area (Sub-areas 6 and 7 and Divisions 8.a,b,d,e), 1985-2017. Continued.

| COUNTRY | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 1 | 2 |  |  |  |  | 14 | 44 | 21 |
| Denmark |  | 48 | 2889 | 8 | 903 | 18538 | 6741 | 19443 | 12569 |
| Estonia |  |  |  |  |  |  |  |  |  |
| Faroe <br> Islands | 4421 | 36 | 8 |  |  | 3421 | 5851 | 13173 | 20559 |
| France | 16464 | 10301 | 11304 | 14448 | 12438 | 16627 | 17820 | 16634 | 16925 |
| Germany | 17545 | 16493 | 18792 | 14277 | 15102 | 23478 | 19238 | 9740 | 9608 |
| Guernsey |  |  | 10 | 5 | 9 | 9 | 4 |  |  |
| Ireland | 48155 | 43355 | 45696 | 42627 | 42988 | 56286 | 54571 | 52087 | 48957 |
| Isle of Man |  | 14 | 11 | 11 | 8 | 3 |  | 8 | 2 |
| Jersey | 8 | 6 | 7 | 8 | 8 | 7 | 3 | 3 | 0.003 |
| Lithuania |  |  | 23 |  |  | 176 | 554 | 13 |  |
| Netherlands | 20900 | 21699 | 18336 | 19794 | 16295 | 16242 | 15264 | 17896 | 18694 |
| Norway | 121 | 30 | 2019 | 1101 | 734 |  | 1313 | 1035 | 2657 |
| Poland |  |  |  |  |  |  |  |  |  |
| Russia |  | 1 |  |  |  |  |  | 30 |  |
| Spain | 8462 | 6532 | 1257 | 773 | 635 | 1796 | 951 | 1253 | 786 |
| United Kingdom | 109147 | 107840 | 111103 | 93775 | 92957 | 137195 | 110932 | 112268 | 116308 |
| Misreported <br> (Area 4.a) | -1959 |  |  |  |  |  |  |  |  |
| Misreported (Unknown) |  |  |  |  |  |  |  |  |  |
| Unallocated | 490 | 4503 | 399 | 16 | -144 |  | 34 |  |  |
| Discards | 6848 | 7518 | 7153 | 10654 | 2105 | 1742 | 3185 | 2126 | 2142 |
| Total | 230603 | 218377 | 219007 | 197496 | 183857 | 275519 | 236475 | 245754 | 249229 |

Table 8.4.2.4. NE Atlantic Mackerel. ICES estimated catch ( $\mathbf{t}$ ) in Divisions 8.c and 9.a, 1977-2017 (Data submitted by Working Group members).

| COUNTRY | DIV | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | 8.c |  |  |  |  |  |  |  |  |  |
| Poland | 9.a | 8 |  |  |  |  |  |  |  |  |
| Portugal | 9.a | 1743 | 1555 | 1071 | 1929 | 3108 | 3018 | 2239 | 2250 | 4178 |
| Spain | 8.c | 19852 | 18543 | 15013 | 11316 | 12834 | 15621 | 10390 | 13852 | 11810 |
| Spain | 9.a | 2935 | 6221 | 6280 | 2719 | 2111 | 2437 | 2224 | 4206 | 2123 |
| USSR | 9.a | 2879 | 189 | 111 |  |  |  |  |  |  |
| Total | 9.a | 7565 | 7965 | 7462 | 4648 | 5219 | 5455 | 4463 | 6456 | 6301 |
| Total |  | 27417 | 26508 | 22475 | 15964 | 18053 | 21076 | 14853 | 20308 | 18111 |
| Country | Div | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| France | 8.c |  |  |  |  |  |  |  |  |  |
| Poland | 9.a |  |  |  |  |  |  |  |  |  |
| Portugal | 9.a | 6419 | 5714 | 4388 | 3112 | 3819 | 2789 | 3576 | 2015 | 2158 |
| Spain | 8.c | 16533 | 15982 | 16844 | 13446 | 16086 | 16940 | 12043 | 16675 | 21246 |
| Spain | 9.a | 1837 | 491 | 3540 | 1763 | 1406 | 1051 | 2427 | 1027 | 1741 |
| USSR | 9.a |  |  |  |  |  |  |  |  |  |
| Total | 9.a | 8256 | 6205 | 7928 | 4875 | 5225 | 3840 | 6003 | 3042 | 3899 |
| Total |  | 24789 | 22187 | 24772 | 18321 | 21311 | 20780 | 18046 | 19719 | 25045 |
| Country | Div | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| France | 8.c |  |  |  |  |  |  |  |  | 226 |
| Poland | 9.a |  |  |  |  |  |  |  |  |  |
| Portugal | 9.a | 2893 | 3023 | 2080 | 2897 | 2002 | 2253 | 3119 | 2934 | 2749 |
| Spain | 8.c | 23631 | 28386 | 35015 | 36174 | 37631 | 30061 | 38205 | 38703 | 17384 |
| Spain | 9.a | 1025 | 2714 | 3613 | 5093 | 4164 | 3760 | 1874 | 7938 | 5464 |
| Discards | 8.c |  |  |  |  |  |  |  |  | 531 |
| Discards | 9.a | 3918 | 5737 | 5693 | 7990 | 6165 | 6013 |  |  |  |
| Total | 9.a | 27549 | 34123 | 40708 | 44164 | 43796 | 36074 | 4993 | 10873 | 8213 |
| Total |  |  |  |  |  |  |  | 43198 | 49575 | 26354 |

Table 8.4.2.4. NE Atlantic Mackerel. ICES estimated catch ( $\mathbf{t}$ ) in Divisions 8.c and 9.a, 1977-2017 (Data submitted by Working Group members). Continued.

| CounTRY | DIV | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| France | 8.c | 177 | 151 | 43 | 55 | 168 | 383 | 392 | 44 | 283 |
| Poland | 9.a |  |  |  |  |  |  |  |  |  |
| Portugal | 9.a | 2289 | 1509 | 2620 | 2605 | 2381 | 1753 | 2363 | 962 | 824 |
| Spain | 8.c |  |  | 43063 | 53401 | 50455 | 91043 | 38858 | 14709 | 17768 |
| Spain | 9.a |  |  | 7025 | 6773 | 6855 | 14569 | 7347 | 2759 | 845 |
| Discards | 8.c | 928 | 391 | 3606 | 156 | 73 | 725 | 4408 | 563 | 2187 |
| Discards | 9.a |  | 405 | 1 | 916 | 677 | 241 | 232 | 1245 | 1244 |
| Unallocated | 8.c | 28429 | 42851 |  |  |  |  |  | 4691 | 4144 |
| Unallocated | 9.a | 3946 | 5107 |  |  |  |  | 108 | 871 | 1076 |
| Total | 9.a | 6234 | 7021 | 9646 | 10293 | 9913 | 16562 | 10049 | 5836 | 3989 |
| Total |  | 35768 | 50414 | 56358 | 63906 | 60609 | 108713 | 53708 | 25843 | 28372 |
|  |  |  |  |  |  |  |  |  |  |  |
| Country | Div | 2013 | 2014 | 2015 | 2016 | 2017 |  |  |  |  |
| France | 8.c | 220 | 171 | 21 | 106 | 83 |  |  |  |  |
| Portugal | 9.a | 254 | 618 | 1456 | 619 | 634 |  |  |  |  |
| Spain | 8.c | 14617 | 33783 | 29726 | 26553 | 30893 |  |  |  |  |
| Spain | 9.a | 1162 | 2227 | 3853 | 2229 | 1206 |  |  |  |  |
| Discards | 8.c | 1428 | 2821 | 4724 | 2469 | 84 |  |  |  |  |
| Discards | 9.a | 1027 | 1463 | 2409 | 751 | 143 |  |  |  |  |
| Unallocated | 8.c | -573 | 8795 | 11 | 1357 |  |  |  |  |  |
| Unallocated | 9.a | 4053 | 662 | 1831 | 2123 |  |  |  |  |  |
| Total | 9.a | 6497 | 4308 | 9550 | 5722 | 1983 |  |  |  |  |
| Total |  | 22188 | 45570 | 44033 | 36207 | 33042 |  |  |  |  |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017.

Quarters 1-4

| Age | 2.a | 2.a. 1 | 2.a. 2 | $3 . a$ | 3.6 | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 6469.5 | 1.2 | 166.3 | 22.5 | 1.2 | 1.1 | 0.1 | 16153.1 | 59.3 |
| 2 | 4144.3 | 2.4 | 332.5 | 51.0 | 1.9 | 1.0 | 0.3 | 47117.9 | 1966.2 |
| 3 | 184206.6 | 40.7 | 1342.7 | 278.4 | 8.3 | 10.9 | 1.6 | 144879.1 | 5758.8 |
| 4 | 137156.9 | 56.4 | 1360.5 | 196.7 | 0.6 | 4.7 | 1.7 | 93088.2 | 1235.3 |
| 5 | 130056.2 | 104.1 | 1511.4 | 224.5 | 2.4 | 8.8 | 1.4 | 96425.7 | 1163.5 |
| 6 | 207130.8 | 274.1 | 1374.7 | 284.7 | 6.2 | 12.3 | 1.5 | 100433.4 | 1241.8 |
| 7 | 192363.7 | 193.2 | 930.2 | 269.3 | 6.2 | 8.1 | 1.1 | 81246.2 | 1142.2 |
| 8 | 98036.7 | 95.1 | 250.1 | 186.5 | 3.5 | 5.8 | 0.6 | 51452.8 | 935.5 |
| 9 | 69092.1 | 50.1 | 483.4 | 104.0 | 2.0 | 2.9 | 0.3 | 31025.3 | 428.7 |
| 10 | 52518.2 | 19.8 | 591.1 | 98.1 | 2.0 | 2.0 | 0.4 | 25330.3 | 550.1 |
| 11 | 21175.0 | 1.0 | 143.3 | 28.2 | 0.6 | 0.6 | 0.1 | 9160.7 | 44.2 |
| 12 | 13075.8 | 1.0 | 143.3 | 23.0 | 0.1 | 0.1 | 0.1 | 5059.8 | 138.8 |
| 13 | 5163.8 |  |  | 16.3 | 0.1 |  | 0.1 | 1819.4 | 129.2 |
| 14 | 1652.8 |  |  | 10.7 | 0.1 | 0.1 |  | 844.7 | 105.8 |
| 15+ | 1328.8 |  |  | 0.8 |  |  |  | 220.1 | 5.4 |
| Catch | 461313.5 | 404.3 | 3636.5 | 686.0 | 12.4 | 22.8 | 3.5 | 263824.6 | 4723.4 |
| SOP | 461303.9 | 404.3 | 3636.5 | 686.3 | 12.4 | 22.8 | 3.6 | 263661.4 | 4725.4 |
| SOP\% | 100\% | 100\% | 100\% | 100\% | 100\% | 100\% | 98\% | 100\% | 100\% |
| AGE | 4.c | 5.a | 5.b | 6.a | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  | 36.1 |  |  |  |  |  |
| 1 | 22.3 |  | 29.6 | 1896.2 |  | 0.1 | 30.3 | 20.3 | 395.9 |
| 2 | 516.6 | 908.6 | 128.2 | 4510.6 | 47.2 | 0.9 | 182.6 | 6.6 | 5676.2 |
| 3 | 436.5 | 8545.6 | 3134.2 | 46897.1 | 106.3 | 2.8 | 1378.1 | 3.3 | 5653.7 |
| 4 | 107.2 | 20461.0 | 2819.6 | 58059.9 | 39.7 | 2.3 | 770.4 | 1.4 | 1499.6 |
| 5 | 207.1 | 39384.5 | 3956.7 | 85304.6 | 0.9 | 4.5 | 2813.3 | 2.7 | 2376.4 |
| 6 | 135.8 | 49629.6 | 5602.4 | 105243.8 | 1.0 | 4.0 | 3360.9 | 5.2 | 1637.2 |
| 7 | 98.1 | 32370.6 | 4854.3 | 94518.4 | 0.7 | 5.5 | 2314.0 | 6.4 | 883.3 |
| 8 | 61.0 | 19226.1 | 3360.7 | 72337.6 | 0.4 | 2.5 | 2933.7 | 3.2 | 719.1 |
| 9 | 58.1 | 15585.2 | 1941.0 | 59501.7 | 0.3 | 3.5 | 2509.9 | 4.7 | 470.2 |
| 10 | 48.3 | 7674.4 | 1771.6 | 44873.8 | 0.2 | 3.1 | 1219.3 | 2.3 | 316.5 |
| 11 | 22.7 | 2953.6 | 222.5 | 23383.1 | 0.1 | 1.6 | 722.1 | 1.4 | 105.8 |
| 12 | 4.5 | 1420.9 | 424.4 | 11126.9 |  | 0.4 | 113.2 |  | 44.2 |
| 13 | 2.6 | 4.9 | 325.3 | 4280.6 |  | 0.1 | 65.0 |  | 17.0 |
| 14 | 2.1 |  | 269.6 | 1803.1 |  | 0.0 | 17.3 |  | 6.9 |
| 15+ | 1.2 |  |  | 600.7 |  | 0.0 | 0.0 |  | 12.6 |
| Catch | 531.6 | 87734.0 | 11343.6 | 226056.0 | 53.6 | 10.5 | 6420.5 | 17.8 | 6082.3 |
| SOP | 528.0 | 87738.8 | 11343.3 | 226102.9 | 54.9 | 10.5 | 6421.6 | 17.8 | 6033.4 |
| SOP\% | 101\% | 100\% | 100\% | 100\% | 97\% | 100\% | 100\% | 100\% | 101\% |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017 (cont.).
Quarters 1-4

| Age | 7.e | 7.f | 7.9 | 7.h | 7.j | 7.k | 8.a | 8.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 10111.7 | 17053.3 |
| 1 | 237.5 | 36.2 | 71.0 | 58.8 | 352.4 | 0.0 | 2086.3 | 6006.8 |
| 2 | 332.5 | 246.3 | 17.4 | 17.4 | 100.1 | 0.1 | 1808.5 | 3365.2 |
| 3 | 835.7 | 1273.6 | 3.7 | 9.1 | 214.9 | 0.1 | 1375.5 | 2963.7 |
| 4 | 360.5 | 868.3 | 1.1 | 2.9 | 193.8 | 0.04 | 632.4 | 1239.8 |
| 5 | 236.9 | 319.0 | 1.3 | 8.1 | 488.7 | 0.05 | 669.0 | 2277.2 |
| 6 | 194.6 | 147.3 | 1.8 | 7.6 | 897.5 | 0.04 | 673.1 | 2187.7 |
| 7 | 150.0 | 79.8 | 1.6 | 4.4 | 902.2 | 0.04 | 485.5 | 1411.7 |
| 8 | 161.4 | 51.4 | 1.2 | 6.7 | 687.0 | 0.01 | 278.8 | 726.4 |
| 9 | 130.9 | 39.1 | 1.1 | 5.3 | 622.2 | 0.02 | 154.5 | 401.7 |
| 10 | 108.8 | 20.4 | 0.5 | 2.8 | 281.0 | 0.01 | 95.9 | 150.9 |
| 11 | 147.5 | 18.1 | 0.3 | 3.0 | 194.8 | 0.01 | 16.6 | 38.4 |
| 12 | 49.0 | 6.6 | 0.1 | 0.6 | 93.0 | 0.00 | 29.8 | 7.3 |
| 13 | 63.9 | 8.8 | 0.1 | 0.7 | 69.8 | 0.0 | 0.9 | 2.1 |
| 14 | 32.0 | 4.4 | 0.0 | 0.3 | 26.5 | 0.0 | 0.0 | 0.0 |
| 15+ | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Catch | 956.0 | 678.7 | 22.5 | 34.8 | 1817.3 | 0.2 | 2149.6 | 4853.5 |
| SOP | 956.7 | 678.6 | 22.5 | 34.8 | 1817.3 | 0.2 | 2154.2 | 4863.2 |
| SOP\% | 100\% | 100\% | 100\% | 100\% | 100\% | 98\% | 100\% | 100\% |
| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.9 | 14.b | All |
| 0 | 329.63 | 0.0 | 59.5 | 435.9 | 279.8 | 0.0 | 0.0 | 28306.0 |
| 1 | 2702.36 | 1032.4 | 47.8 | 1319.9 | 1370.4 | 13.5 | 2853.7 | 43457.9 |
| 2 | 4991.5 | 1605.4 | 18.5 | 346.3 | 1019.5 | 38.7 | 8236.9 | 87739.1 |
| 3 | 8591.24 | 1919.1 | 18.8 | 811.0 | 1055.6 | 170.5 | 36373.9 | 458301.2 |
| 4 | 5192.66 | 2098.1 | 10.9 | 338.9 | 462.5 | 107.8 | 23407.2 | 351779.0 |
| 5 | 11664.52 | 5844.9 | 43.8 | 93.4 | 403.8 | 47.9 | 11215.3 | 396862.3 |
| 6 | 13062.37 | 6765.4 | 38.2 | 66.0 | 356.3 | 7.0 | 2816.8 | 503601.1 |
| 7 | 10188.69 | 5306.6 | 42.5 | 65.8 | 243.5 | 0.2 | 920.3 | 431014.2 |
| 8 | 6487.7 | 3208.7 | 15.1 | 47.2 | 146.5 | 0.0 | 530.5 | 261959.5 |
| 9 | 3844.51 | 1899.4 | 21.9 | 32.8 | 91.5 | 0.0 | 441.8 | 188949.9 |
| 10 | 1444.85 | 694.8 | 14.3 | 51.8 | 38.7 | 0.0 | 217.0 | 138143.1 |
| 11 | 446.91 | 265.3 | 8.5 | 4.7 | 14.5 | 0.0 | 85.9 | 59210.9 |
| 12 | 154.2 | 40.6 | 0.2 | 19.2 | 4.8 | 0.0 | 40.4 | 32022.0 |
| 13 | 94.33 | 53.5 | 0.1 | 0.0 | 3.0 | 0.0 | 0.0 | 12121.5 |
| 14 | 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 4776.4 |
| 15+ | 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2169.6 |
| Catch | 20769.4 | 10289.9 | 75.9 | 776.6 | 1206.1 | 173.6 | 39263.1 | 1155943.8 |
| SOP | 20767.8 | 10289.5 | 75.9 | 776.6 | 1206.2 | 173.5 | 39263.9 | 1155785.1 |
| SOP\% | 100\% | 100\% | 100\% | 100\% | 100\% | 100\% | 100\% | 100\% |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017 (cont.).

## Quarter 1

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | 3.b | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  | 3.8 |  |
| 2 |  |  |  |  |  |  |  | 36.1 |  |
| 3 | 0.6 |  |  | 1.3 |  |  | 0.1 | 4832.1 | 0.1 |
| 4 | 0.6 |  |  | 0.8 |  |  |  | 2996.4 |  |
| 5 | 1.3 |  |  | 0.7 |  |  |  | 2655.8 | 0.1 |
| 6 | 1.9 |  |  | 1.1 |  |  | 0.1 | 4409.7 | 0.1 |
| 7 | 1.5 |  |  | 1.5 |  |  | 0.1 | 5380.1 | 0.1 |
| 8 | 1.0 |  |  | 1.2 |  |  | 0.1 | 4235.3 | 0.1 |
| 9 | 0.7 |  |  | 0.5 |  |  |  | 1846.4 |  |
| 10 | 0.5 |  |  | 0.7 |  |  |  | 2557.5 |  |
| 11 | 0.2 |  |  |  |  |  |  | 81.2 |  |
| 12 | 0.1 |  |  | 0.2 |  |  |  | 680.3 |  |
| 13 | 0.1 |  |  | 0.2 |  |  |  | 661.7 |  |
| 14 |  |  |  | 0.2 |  |  |  | 547.6 |  |
| 15+ |  |  |  |  |  |  |  | 0.0 |  |
| Catch | 3.74 |  |  | 2.85 |  |  | 0.13 | 10717.97 | 0.13 |
| SOP | 3.74 |  |  | 2.87 |  |  | 0.13 | 10720.48 | 0.13 |
| SOP\% | 100\% |  |  | 99\% |  |  | 102\% | 100\% | 99\% |
| AGE | 4.c | 5.a | 5.b | 6.a | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  | 35.9 |  |  |  |  |  |
| 1 |  |  |  | 1871.0 |  |  | 30.2 | 20.3 | 92.2 |
| 2 | 0.7 |  | 0.1 | 4255.7 |  |  | 170.2 | 4.9 | 73.7 |
| 3 | 7.1 | 0.1 | 2156.5 | 46521.3 |  | 0.6 | 1292.3 | 1.0 | 571.0 |
| 4 | 4.3 | 0.3 | 1347.8 | 57870.8 |  | 0.8 | 726.9 | 0.8 | 337.4 |
| 5 | 7.1 | 0.5 | 1078.3 | 83936.6 |  | 1.5 | 2645.6 | 1.9 | 560.4 |
| 6 | 6.3 | 0.6 | 1886.9 | 104111.4 |  | 2.1 | 3178.4 | 4.7 | 516.8 |
| 7 | 5.5 | 0.4 | 2426.0 | 92769.9 |  | 2.6 | 2184.3 | 6.2 | 447.9 |
| 8 | 3.6 | 0.2 | 1940.8 | 71779.2 |  | 1.7 | 2750.0 | 3.2 | 296.4 |
| 9 | 2.7 | 0.1 | 808.7 | 58446.8 |  | 1.5 | 2365.1 | 4.6 | 213.5 |
| 10 | 1.2 | 0.1 | 1186.0 | 44031.6 |  | 1.6 | 1144.4 | 2.3 | 102.3 |
| 11 | 0.1 |  | 0.0 | 22776.0 |  | 0.7 | 630.3 | 1.3 | 10.9 |
| 12 | 0.2 |  | 323.5 | 11049.4 |  | 0.4 | 90.7 |  | 20.1 |
| 13 | 0.1 |  | 323.5 | 4214.6 |  | 0.1 | 38.0 |  | 11.2 |
| 14 | 0.0 |  | 269.6 | 1769.9 |  | 0.0 | 4.3 |  | 1.2 |
| 15+ | 0.0 |  | 0.0 | 598.3 |  | 0.0 | 0.0 |  | 0.0 |
| Catch | 13.1 | 1.0 | 4765.1 | 223122.1 |  | 5.1 | 5974.1 | 15.6 | 1082.3 |
| SOP | 13.1 | 1.0 | 4765.3 | 223167.9 |  | 5.1 | 5973.9 | 15.6 | 1082.6 |
| SOP\% | 100\% | 99\% | 100\% | 100\% |  | 100\% | 100\% | 100\% | 100\% |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017 (cont.).

Quarter 1

| Age | $7 . \mathrm{e}$ | 7.f | 7.9 | 7.h | 7.j | 7.k | 8.a | 8.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  | 1811.5 | 4297.5 |
| 1 | 208.7 | 10.1 | 71.0 | 58.8 | 352.4 |  | 776.3 | 4083.8 |
| 2 | 66.0 | 52.8 | 16.2 | 13.8 | 91.1 |  | 363.7 | 2585.9 |
| 3 | 146.3 | 649.2 | 1.3 | 3.9 | 190.6 |  | 331.8 | 2363.1 |
| 4 | 93.0 | 464.2 | 0.1 | 1.4 | 181.7 |  | 132.9 | 904.0 |
| 5 | 85.4 | 137.5 | 0.1 | 5.7 | 461.6 |  | 267.9 | 1833.3 |
| 6 | 75.2 | 53.0 | 0.1 | 5.6 | 859.5 | 0.020 | 248.0 | 1716.4 |
| 7 | 44.0 | 21.7 | 0.1 | 3.0 | 836.3 | 0.030 | 147.4 | 1037.8 |
| 8 | 57.5 | 9.7 | 0.1 | 5.3 | 623.3 | 0.010 | 70.1 | 503.5 |
| 9 | 45.6 | 10.7 | 0.1 | 4.4 | 584.7 | 0.020 | 36.9 | 266.4 |
| 10 | 25.3 | 2.8 | 0.0 | 2.2 | 257.5 | 0.010 | 12.8 | 92.5 |
| 11 | 14.5 | 0.6 | 0.0 | 1.4 | 96.5 | 0.010 | 3.2 | 23.2 |
| 12 | 1.2 |  |  |  | 56.4 |  | 0.3 | 2.4 |
| 13 | 0.4 |  |  |  | 22.7 |  | 0.2 | 1.3 |
| 14 | 0.3 |  |  |  | 3.2 |  | 0.0 | 0.0 |
| 15+ | 0.0 |  |  |  |  |  | 0.0 | 0.0 |
| Catch | 228.5 | 281.5 | 18.4 | 26.1 | 1570.7 | 0.04 | 509.2 | 3464.5 |
| SOP | 228.6 | 281.5 | 18.4 | 26.1 | 1570.8 | 0.04 | 510.1 | 3470.5 |
| SOP\% | 100\% | 100\% | 100\% | 100\% | 100\% | 93\% | 100\% | 100\% |
| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | $14 . a$ | 14.b | All |
| 0 | 0 | 0 | 5.53 | 0 | 0 |  |  | 6150.4 |
| 1 | 1560.4 | 1026.0 | 25.0 | 66.9 | 0.4 |  |  | 10257.2 |
| 2 | 3533.4 | 1501.7 | 12.9 | 87.1 | 25.1 |  |  | 12891.2 |
| 3 | 4900.4 | 1685.8 | 17.3 | 154.6 | 164.1 |  |  | 65992.4 |
| 4 | 2974.0 | 1828.4 | 10.1 | 132.1 | 118.9 |  |  | 70127.6 |
| 5 | 6847.1 | 5110.5 | 22.8 | 12.6 | 194.5 |  |  | 105868.6 |
| 6 | 7487.7 | 5906.1 | 25.3 | 15.1 | 161.4 |  |  | 130673.6 |
| 7 | 5445.8 | 4603.0 | 18.5 | 35.2 | 69.3 |  |  | 115488.0 |
| 8 | 3203.8 | 2785.9 | 11.0 | 23.6 | 26.0 |  |  | 88332.2 |
| 9 | 1820.3 | 1620.0 | 6.2 | 18.9 | 12.2 |  |  | 68117.0 |
| 10 | 670.5 | 580.8 | 2.5 | 40.6 | 4.7 |  |  | 50720.1 |
| 11 | 172.5 | 188.1 | 0.7 | 2.7 | 1.6 |  |  | 24005.7 |
| 12 | 50.4 | 36.7 | 0.2 | 19.2 | 1.0 |  |  | 12332.8 |
| 13 | 12.3 | 15.0 | 0.1 |  | 0.4 |  |  | 5301.8 |
| 14 |  |  | 0.0 |  |  |  |  | 2596.3 |
| 15+ |  |  |  |  |  |  |  | 598.4 |
| Catch | 11357.5 | 8906.9 | 39.9 | 191.0 | 216.7 |  |  | 272514.0 |
| SOP | 11357.2 | 8906.9 | 39.9 | 191.0 | 216.7 |  |  | 272586.7 |
| SOP\% | 100\% | 100\% | 100\% | 100\% | 100\% |  |  | 100\% |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017 (cont.).

## Quarter 2

| Age | 2.a | 2.a.1 | 2.a.2 | 3.a | 3.b | 3.c | 3.d | 4.a | $4 . \mathbf{b}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 56.3 |  |  | 1.2 |  |  |  | 6.4 | 18.5 |
| 2 | 77.2 |  |  | 0.0 |  |  |  | 55.1 | 1759.6 |
| 3 | 12817.9 |  |  |  |  |  |  |  |  |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017 (cont.).

Quarter 2

| Age | 7.e | 7.f | 7.9 | 7.h | 7.j | 7.k | $8 . a$ | 8.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1082.6 | 3639.6 |
| 1 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 457.3 | 1545.1 |
| 2 | 43.2 | 5.4 | 0.6 | 3.5 | 6.4 | 0.1 | 63.5 | 141.5 |
| 3 | 104.6 | 27.4 | 1.1 | 5.1 | 18.2 | 0.1 | 106.2 | 178.9 |
| 4 | 42.3 | 17.1 | 0.5 | 1.5 | 9.9 | 0.0 | 104.6 | 148.7 |
| 5 | 49.1 | 9.3 | 1.1 | 2.5 | 27.1 | 0.0 | 299.8 | 395.2 |
| 6 | 40.8 | 8.0 | 1.5 | 2.0 | 38.0 | 0.0 | 349.3 | 440.6 |
| 7 | 19.8 | 3.2 | 1.3 | 1.0 | 65.6 | 0.0 | 299.2 | 358.7 |
| 8 | 25.7 | 5.4 | 1.0 | 0.9 | 63.4 | 0.0 | 183.8 | 214.1 |
| 9 | 20.4 | 4.6 | 0.9 | 0.8 | 37.4 | 0.0 | 117.6 | 135.3 |
| 10 | 11.0 | 5.7 | 0.4 | 0.4 | 23.4 | 0.0 | 44.6 | 51.1 |
| 11 | 6.3 | 0.1 | 0.2 | 0.2 | 97.5 | 0.0 | 13.4 | 15.2 |
| 12 | 0.3 | 0.0 | 0.1 | 0.0 | 36.3 | 0.0 | 1.4 | 1.6 |
| 13 | 0.1 | 0.0 | 0.0 | 0.0 | 46.7 | 0.0 | 0.8 | 0.9 |
| 14 | 0.1 | 0.0 | 0.0 | 0.0 | 23.1 | 0.0 | 0.0 | 0.0 |
| 15+ | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Catch | 112.0 | 22.1 | 3.1 | 6.2 | 242.2 | 0.1 | 596.7 | 815.5 |
| SOP | 112.0 | 22.1 | 3.1 | 6.2 | 242.2 | 0.1 | 600.2 | 818.6 |
| SOP\% | 100\% | 100\% | 100\% | 100\% | 100\% | 102\% | 99\% | 100\% |
| Age | 8.c | 8.c.E | 8.d | $9 . \mathrm{a}$ | 9.a.N | 14.9 | $14 . \mathrm{b}$ | All |
| 0 | 0.0 | 0.0 | 54.0 | 0.0 | 0.0 |  | 0.0 | 4776.4 |
| 1 | 1025.8 | 2.2 | 22.7 | 863.2 | 59.7 |  | 6.0 | 4120.5 |
| 2 | 1335.2 | 52.6 | 5.5 | 43.4 | 377.3 |  | 17.2 | 5412.3 |
| 3 | 3522.2 | 165.4 | 1.5 | 333.1 | 353.1 |  | 75.7 | 23780.1 |
| 4 | 2165.3 | 243.2 | 0.8 | 107.3 | 94.0 |  | 47.8 | 14675.3 |
| 5 | 4806.9 | 729.2 | 21.0 | 56.4 | 147.7 |  | 21.3 | 15913.8 |
| 6 | 5570.1 | 856.7 | 13.0 | 46.4 | 178.8 |  | 3.1 | 17876.8 |
| 7 | 4740.2 | 702.5 | 24.1 | 28.3 | 167.6 |  | 0.1 | 18157.8 |
| 8 | 3282.6 | 421.8 | 4.2 | 23.5 | 119.5 |  | 0.0 | 8538.2 |
| 9 | 2024.2 | 279.4 | 15.7 | 13.9 | 79.3 |  | 0.0 | 5997.5 |
| 10 | 773.8 | 111.3 | 11.7 | 11.3 | 34.0 |  | 0.0 | 3599.8 |
| 11 | 274.4 | 77.2 | 7.8 | 2.0 | 12.8 |  | 0.0 | 1558.6 |
| 12 | 103.7 | 3.8 | 0.0 | 0.0 | 3.8 |  | 0.0 | 845.0 |
| 13 | 82.1 | 38.5 | 0.0 | 0.0 | 2.6 |  | 0.0 | 394.6 |
| 14 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  | 0.0 | 168.4 |
| 15+ | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  | 0.0 | 24.6 |
| Catch | 9289.1 | 1338.4 | 36.0 | 272.9 | 437.2 |  | 77.0 | 39972.4 |
| SOP | 9289.3 | 1338.5 | 36.0 | 272.9 | 437.1 |  | 77.0 | 39971.2 |
| SOP\% | 100\% | 100\% | 100\% | 100\% | 100\% |  | 100\% | 100\% |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017 (cont.).

## Quarter 3

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | 3.b | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 940.1 |  |  | 17.4 | 1.2 | 1.0 | 0.1 | 730.6 | 30.4 |
| 2 | 3076.3 |  |  | 40.3 | 1.9 | 0.9 | 0.2 | 1249.7 | 171.7 |
| 3 | 139839.6 | 31.3 | 35.6 | 150.7 | 7.8 | 9.7 | 1.1 | 4961.8 | 1657.7 |
| 4 | 115443.9 | 47.0 | 53.4 | 110.1 | 0.4 | 4.2 | 1.3 | 1571.0 | 395.3 |
| 5 | 107814.9 | 94.0 | 106.8 | 131.0 | 2.2 | 7.8 | 1.0 | 1899.1 | 262.8 |
| 6 | 170535.3 | 266.4 | 302.6 | 147.5 | 5.9 | 11.0 | 1.0 | 2730.8 | 192.5 |
| 7 | 155562.2 | 188.0 | 213.6 | 120.4 | 5.7 | 7.2 | 0.6 | 2482.0 | 83.0 |
| 8 | 75970.7 | 94.0 | 106.8 | 79.1 | 3.1 | 5.2 | 0.3 | 1754.1 | 37.2 |
| 9 | 51287.8 | 47.0 | 53.4 | 45.9 | 1.8 | 2.6 | 0.2 | 1405.0 | 19.7 |
| 10 | 37407.1 | 15.7 | 17.8 | 35.0 | 1.8 | 1.8 | 0.1 | 1176.3 | 13.5 |
| 11 | 14848.1 |  |  | 13.4 | 0.6 | 0.5 | 0.1 | 507.1 | 3.8 |
| 12 | 9157.3 |  |  | 6.4 |  | 0.1 |  | 280.9 | 2.0 |
| 13 | 4390.8 |  |  | 1.7 |  | 0.0 |  | 65.4 | 1.1 |
| 14 | 1470.4 |  |  | 0.5 |  | 0.1 |  | 15.7 | 1.0 |
| 15+ | 1127.5 |  |  | 0.2 |  |  |  | 19.2 | 0.2 |
| Catch | 369931.2 | 381.3 | 433.2 | 350.7 | 11.5 | 20.3 | 2.3 | 7734.3 | 911.4 |
| SOP | 369913.5 | 381.3 | 433.2 | 350.8 | 11.5 | 20.3 | 2.4 | 7735.0 | 912.8 |
| SOP\% | 100\% | 100\% | 100\% | 100\% | 100\% | 100\% | 96\% | 100\% | 100\% |
| AGE | 4.c | 5.a | 5.b | 6.a | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 10.4 |  |  |  |  |  | 0.1 |  | 124.2 |
| 2 | 230.8 | 883.7 | 57.7 | 1.4 | 47.1 | 0.7 | 0.3 | 0.06 | 2530.6 |
| 3 | 192.9 | 8372.0 | 546.0 | 0.4 | 106.0 | 1.1 | 1.1 | 0.14 | 2174.9 |
| 4 | 45.2 | 20121.7 | 1311.8 | 0.3 | 39.6 | 0.9 | 0.9 | 0.05 | 492.7 |
| 5 | 73.4 | 38720.3 | 2524.1 | 6.6 | 0.8 | 2.6 | 1.0 | 0.00 | 799.8 |
| 6 | 45.3 | 48843.8 | 3184.0 | 4.1 | 1.0 | 1.5 | 1.1 | 0.00 | 489.2 |
| 7 | 17.5 | 31904.1 | 2079.8 | 7.7 | 0.6 | 2.7 | 9.9 | 0.01 | 179.0 |
| 8 | 18.2 | 18969.4 | 1236.6 | 1.4 | 0.4 | 0.5 | 9.6 |  | 186.8 |
| 9 | 10.5 | 15428.0 | 1005.7 | 5.1 | 0.3 | 1.8 | 3.3 |  | 108.4 |
| 10 | 8.8 | 7596.8 | 495.2 | 3.8 | 0.2 | 1.3 | 3.2 |  | 90.4 |
| 11 | 4.0 | 2933.3 | 191.2 | 2.5 | 0.1 | 0.9 | 24.2 |  | 40.9 |
| 12 | 1.1 | 1407.6 | 91.8 |  |  |  | 9.1 |  | 10.8 |
| 13 | 0.3 | 4.8 | 0.3 |  |  |  | 12.1 |  | 2.5 |
| 14 | 0.2 |  |  |  |  |  | 6.0 |  | 2.5 |
| 15+ | 0.5 |  |  |  |  |  |  |  | 5.7 |
| Catch | 200.1 | 86476.0 | 5637.2 | 11.1 | 53.4 | 4.4 | 47.2 | 0.1 | 2187.0 |
| SOP | 198.7 | 86476.7 | 5637.3 | 11.1 | 54.7 | 4.4 | 47.2 | 0.1 | 2164.3 |
| SOP\% | 101\% | 100\% | 100\% | 100\% | 98\% | 100\% | 100\% | 104\% | 101\% |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017 (cont.).

Quarter 3

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | $8 . \mathrm{a}$ | 8.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  | 4520.1 | 2833.5 |
| 1 | 4.9 | 0.5 |  |  |  | 0.00 | 69.5 | 39.2 |
| 2 | 69.9 | 31.6 | 0.5 |  | 2.6 | 0.00 | 457.7 | 128.8 |
| 3 | 183.3 | 159.8 | 1.2 |  | 5.9 | 0.00 | 656.6 | 141.3 |
| 4 | 92.3 | 132.2 | 0.5 |  | 2.2 | 0.01 | 304.5 | 67.0 |
| 5 | 38.4 | 56.1 | 0.0 |  | 0.0 | 0.00 | 83.2 | 18.3 |
| 6 | 23.5 | 26.6 | 0.0 |  | 0.0 | 0.00 | 65.4 | 12.4 |
| 7 | 39.0 | 15.8 | 0.1 | 0.1 | 0.3 | 0.00 | 35.2 | 6.5 |
| 8 | 35.0 | 10.9 | 0.04 | 0.1 | 0.3 | 0.00 | 22.8 | 3.9 |
| 9 | 14.7 | 7.6 | 0.01 | 0.0 | 0.1 | 0.00 | 0.0 | 0.0 |
| 10 | 14.0 | 3.1 | 0.0 | 0.0 | 0.1 | 0.00 | 37.3 | 4.4 |
| 11 | 68.4 | 3.4 | 0.1 | 0.3 | 0.7 | 0.00 | 0.0 | 0.0 |
| 12 | 25.7 | 1.3 |  | 0.1 | 0.3 | 0.00 | 27.8 | 2.7 |
| 13 | 34.2 | 1.7 |  | 0.1 | 0.3 | 0.00 | 0.0 | 0.0 |
| 14 | 17.1 | 0.9 |  | 0.1 | 0.2 | 0.00 | 0.0 | 0.0 |
| 15+ | 0.0 | 0.0 |  | 0.0 | 0.0 | 0.00 | 0.0 | 0.0 |
| Catch | 241.4 | 117.4 | 0.8 | 0.6 | 4.2 | 0.004 | 602.6 | 168.6 |
| SOP | 241.5 | 117.4 | 0.8 | 0.6 | 4.2 | 0.004 | 602.4 | 168.5 |
| SOP\% | 100\% | 100\% | 100\% | 99\% | 100\% | 100\% | 100\% | 100\% |
| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | $14 . \mathrm{b}$ | All |
| 0 | 10.8 |  |  | 280.3 | 245.0 |  |  | 7889.6 |
| 1 | 32.8 |  |  | 264.0 | 163.8 | 13.5 | 2847.7 | 5291.3 |
| 2 | 44.6 | 1.88 | 0.01 | 143.1 | 152.5 | 38.7 | 8219.7 | 17584.8 |
| 3 | 94.3 | 3.87 | 0.06 | 208.3 | 252.2 | 170.5 | 36298.3 | 196265.5 |
| 4 | 35.5 | 2.03 | 0.02 | 48.6 | 126.7 | 107.8 | 23359.4 | 163918.4 |
| 5 | 8.8 | 0.52 | 0.01 | 17.4 | 40.5 | 47.9 | 11194.0 | 163953.4 |
| 6 | 3.9 | 0.25 | 0.01 |  | 9.3 | 7.0 | 2813.7 | 229724.9 |
| 7 | 2.3 | 0.09 | 0.01 |  | 4.7 | 0.2 | 920.2 | 193888.3 |
| 8 | 1.2 | 0.05 |  |  | 0.8 |  | 530.5 | 99078.8 |
| 9 | 0.0 | 0.00 |  |  |  |  | 441.8 | 69890.5 |
| 10 | 0.5 | 0.04 |  |  |  |  | 217.0 | 47145.2 |
| 11 | 0.0 |  |  |  |  |  | 85.9 | 18729.3 |
| 12 | 0.1 |  |  |  |  |  | 40.4 | 11065.4 |
| 13 |  |  |  |  |  |  |  | 4515.4 |
| 14 |  |  |  |  |  |  |  | 1514.7 |
| 15+ |  |  |  |  |  |  |  | 1153.3 |
| Catch | 57.7 | 2.6 | 0.04 | 197.4 | 199.9 | 173.6 | 39186.1 | 515345.6 |
| SOP | 57.7 | 2.6 | 0.04 | 197.4 | 200.1 | 173.5 | 39187.1 | 515307.4 |
| SOP\% | 100\% | 100\% | 97\% | 100\% | 100\% | 100\% | 100\% | 100\% |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017 (cont.).

## Quarter 4

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | $3 . \mathrm{b}$ | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 5473.0 | 1.2 | 166.3 | 4.0 |  | 0.1 | 0.0 | 15412.2 | 10.5 |
| 2 | 990.8 | 2.4 | 332.5 | 10.6 | 0.01 | 0.1 | 0.1 | 45777.0 | 34.9 |
| 3 | 31548.4 | 9.4 | 1307.1 | 39.9 | 0.05 | 1.2 | 0.1 | 134762.0 | 123.2 |
| 4 | 11899.9 | 9.4 | 1307.1 | 33.4 | 0.01 | 0.5 | 0.2 | 88416.1 | 45.8 |
| 5 | 16920.4 | 10.1 | 1404.6 | 38.7 | 0.02 | 0.9 | 0.2 | 91600.2 | 36.9 |
| 6 | 30519.0 | 7.7 | 1072.1 | 40.8 | 0.04 | 1.3 | 0.1 | 92768.0 | 24.9 |
| 7 | 29312.1 | 5.2 | 716.6 | 31.1 | 0.03 | 0.9 | 0.1 | 72814.2 | 12.9 |
| 8 | 20337.7 | 1.0 | 143.3 | 19.9 | 0.02 | 0.6 | 0.1 | 45076.5 | 4.8 |
| 9 | 16769.4 | 3.1 | 430.0 | 10.2 | 0.01 | 0.3 | 0.03 | 27459.9 | 3.0 |
| 10 | 14473.8 | 4.1 | 573.3 | 7.4 | 0.01 | 0.2 | 0.02 | 21356.7 | 1.7 |
| 11 | 6051.6 | 1.0 | 143.3 | 2.6 |  | 0.1 | 0.01 | 8461.7 | 0.7 |
| 12 | 3519.6 | 1.0 | 143.3 | 1.1 |  |  |  | 4028.6 | 0.3 |
| 13 | 761.6 |  |  | 0.3 |  |  |  | 1049.2 | 0.1 |
| 14 | 178.3 |  |  | 0.1 |  |  |  | 268.3 | 0.0 |
| 15+ | 200.4 |  |  |  |  |  |  | 187.9 | 0.1 |
| Catch | 75784.6 | 23.0 | 3203.3 | 93.7 | 0.1 | 2.4 | 0.4 | 244037.9 | 97.3 |
| SOP | 75784.7 | 23.0 | 3203.2 | 93.8 | 0.1 | 2.5 | 0.4 | 243858.9 | 97.6 |
| SOP\% | 100\% | 100\% | 100\% | 100\% | 97\% | 100\% | 99\% | 100\% | 100\% |
| AGE | 4.c | 5.a | 5.b | 6.a | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 1.27 |  | 29.62 | 22.92 |  | 0.02 | 0.01 |  | 136.39 |
| 2 | 55.37 | 0.92 | 44.64 | 35.58 |  | 0.08 | 0.02 |  | 2158.27 |
| 3 | 36.38 | 16 | 190.44 | 150.76 |  | 0.3 | 0.09 |  | 2137.51 |
| 4 | 7.62 | 8.19 | 0.19 | 19.46 |  | 0.17 | 0.07 |  | 494.93 |
| 5 | 15.41 | 16.18 | 47.29 | 53.19 |  | 0.14 | 0.08 |  | 719.08 |
| 6 | 9.33 | 20.67 | 141.2 | 113.42 |  | 0.14 | 0.09 |  | 449.15 |
| 7 | 2.98 | 10.49 | 138.72 | 147.6 |  | 0.09 | 3.51 |  | 178.78 |
| 8 | 4.12 | 4.73 | 75.45 | 95.5 |  | 0.06 | 3.49 |  | 165.42 |
| 9 | 2.36 | 3.48 | 44.44 | 50.62 |  | 0.03 | 1.17 |  | 100.2 |
| 10 | 2.07 | 2.16 | 44.43 | 52.63 |  | 0.02 | 1.16 |  | 83.99 |
| 11 | 0.94 | 0.65 | 14.82 | 122.02 |  | 0 | 9.2 |  | 34.94 |
| 12 | 0.25 | 0.45 |  | 43.17 |  | 0 | 3.45 |  | 9.08 |
| 13 | 0.06 | 0.11 |  | 54.86 |  | 0 | 4.6 |  | 2.12 |
| 14 | 0.06 |  |  | 27.43 |  | 0 | 2.3 |  | 2.12 |
| 15+ | 0.13 |  |  |  |  | 0 | 0 |  | 4.79 |
| Catch | 41.7 | 35.0 | 270.9 | 437.0 |  | 0.3 | 17.2 |  | 2012.2 |
| SOP | 41.6 | 35.0 | 270.9 | 437.0 |  | 0.3 | 17.2 |  | 1993.9 |
| SOP\% | 100\% | 100\% | 100\% | 100\% |  | 100\% | 100\% |  | 101\% |

Table 8.4.3.1. NE Atlantic Mackerel. Catch numbers (‘000s) -at-age by area for 2017 (cont.).

## Quarter 4

| Age | $7 . \mathrm{e}$ | 7.f | 7.9 | 7.h | 7.j | 7.k | 8.a | 8.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  | 2697.5 | 6282.8 |
| 1 | 23.9 | 25.4 | 0.01 |  |  |  | 783.2 | 338.7 |
| 2 | 153.5 | 156.5 | 0.01 |  | 0.06 |  | 923.6 | 509.0 |
| 3 | 401.6 | 437.2 | 0.1 | 0.1 | 0.14 |  | 280.9 | 280.5 |
| 4 | 133.0 | 254.8 | 0.1 | 0.0 | 0.05 |  | 90.6 | 120.1 |
| 5 | 64.0 | 116.1 | 0.1 | 0.0 | 0.00 |  | 18.1 | 30.4 |
| 6 | 55.0 | 59.7 | 0.1 | 0.0 | 0.00 |  | 10.4 | 18.3 |
| 7 | 47.2 | 39.1 | 0.1 | 0.4 | 0.03 |  | 3.8 | 8.8 |
| 8 | 43.2 | 25.4 | 0.1 | 0.4 | 0.03 |  | 2.1 | 4.9 |
| 9 | 50.3 | 16.2 | 0.03 | 0.1 | 0.01 |  | 0.0 | 0.0 |
| 10 | 58.5 | 8.9 | 0.02 | 0.1 | 0.01 |  | 1.2 | 2.9 |
| 11 | 58.4 | 14.1 | 0.02 | 1.1 | 0.07 |  | 0.0 | 0.0 |
| 12 | 21.9 | 5.3 | 0.01 | 0.4 | 0.03 |  | 0.3 | 0.6 |
| 13 | 29.2 | 7.0 | 0.01 | 0.5 | 0.04 |  | 0.0 | 0.0 |
| 14 | 14.6 | 3.5 | 0.00 | 0.3 | 0.02 |  | 0.0 | 0.0 |
| 15+ |  |  |  |  |  |  | 0.0 | 0.0 |
| Catch | 374.1 | 257.7 | 0.2 | 2.0 | 0.2 |  | 441.2 | 404.9 |
| SOP | 374.6 | 257.7 | 0.2 | 2.0 | 0.2 |  | 441.0 | 405.2 |
| SOP\% | 100\% | 100\% | 98\% | 100\% | 95\% |  | 100\% | 100\% |
| Age | 8.c | 8.c.E | 8.d | 9.9 | 9.a.N | $14 . \mathrm{a}$ | $14 . \mathrm{b}$ | All |
| 0 | 318.9 | 0.0 |  | 155.7 | 34.8 |  |  | 9489.6 |
| 1 | 83.5 | 4.2 |  | 125.9 | 1146.5 |  |  | 23788.8 |
| 2 | 78.3 | 49.2 |  | 72.8 | 464.6 |  |  | 51850.8 |
| 3 | 74.4 | 64.0 |  | 115.1 | 286.2 |  |  | 172263.1 |
| 4 | 17.9 | 24.5 |  | 50.9 | 122.9 |  |  | 103057.7 |
| 5 | 1.7 | 4.7 |  | 7.0 | 21.1 |  |  | 111126.5 |
| 6 | 0.7 | 2.4 |  | 4.5 | 6.8 |  |  | 125325.8 |
| 7 | 0.4 | 1.0 |  | 2.3 | 1.9 |  |  | 103480.1 |
| 8 | 0.1 | 1.0 |  | 0.1 | 0.3 |  |  | 66010.3 |
| 9 | 0.0 | 0.0 |  |  |  |  |  | 44944.8 |
| 10 | 0.1 | 2.7 |  |  |  |  |  | 36678.0 |
| 11 |  |  |  |  |  |  |  | 14917.2 |
| 12 |  |  |  |  |  |  |  | 7778.8 |
| 13 |  |  |  |  |  |  |  | 1909.7 |
| 14 |  |  |  |  |  |  |  | 497.1 |
| 15+ |  |  |  |  |  |  |  | 393.3 |
| Catch | 65.0 | 42.0 |  | 115.3 | 352.4 |  |  | 328111.6 |
| SOP | 65.1 | 42.0 |  | 115.3 | 352.4 |  |  | 327929.8 |
| SOP\% | 100\% | 100\% |  | 100\% | 100\% |  |  | 100\% |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<\mathbf{1 \%}$.

Quarters 1-4

| Age | 2.a | 2.a.1 | 2.a.2 | 3.a | 3.b | 3.c | 3.d | 4.a | 4.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | $1 \%$ | $0 \%$ | $2 \%$ | $1 \%$ | $3 \%$ | $2 \%$ | $2 \%$ | $2 \%$ | $0 \%$ |
| 2 | $0 \%$ | $0 \%$ | $4 \%$ | $3 \%$ | $5 \%$ | $2 \%$ | $3 \%$ | $7 \%$ | $13 \%$ |
| 3 | $16 \%$ | $5 \%$ | $16 \%$ | $16 \%$ | $24 \%$ | $19 \%$ | $17 \%$ | $21 \%$ | $39 \%$ |
| 4 | $12 \%$ | $7 \%$ | $16 \%$ | $11 \%$ | $2 \%$ | $8 \%$ | $19 \%$ | $13 \%$ | $8 \%$ |
| 5 | $12 \%$ | $12 \%$ | $18 \%$ | $13 \%$ | $7 \%$ | $15 \%$ | $15 \%$ | $14 \%$ | $8 \%$ |
| 6 | $18 \%$ | $33 \%$ | $16 \%$ | $16 \%$ | $18 \%$ | $21 \%$ | $16 \%$ | $14 \%$ | $8 \%$ |
| 7 | $17 \%$ | $23 \%$ | $11 \%$ | $15 \%$ | $18 \%$ | $14 \%$ | $12 \%$ | $12 \%$ | $8 \%$ |
| 8 | $9 \%$ | $11 \%$ | $3 \%$ | $10 \%$ | $10 \%$ | $10 \%$ | $7 \%$ | $7 \%$ | $6 \%$ |
| 9 | $6 \%$ | $6 \%$ | $6 \%$ | $6 \%$ | $6 \%$ | $5 \%$ | $4 \%$ | $4 \%$ | $3 \%$ |
| 10 | $5 \%$ | $2 \%$ | $7 \%$ | $5 \%$ | $6 \%$ | $3 \%$ | $4 \%$ | $4 \%$ | $4 \%$ |
| 11 | $2 \%$ | $0 \%$ | $2 \%$ | $2 \%$ | $2 \%$ | $1 \%$ | $1 \%$ | $1 \%$ | $0 \%$ |
| 12 | $1 \%$ | $0 \%$ | $2 \%$ | $1 \%$ | $0 \%$ | $0 \%$ | $1 \%$ | $1 \%$ | $1 \%$ |
| 13 | $0 \%$ |  |  | $1 \%$ | $0 \%$ | $0 \%$ | $1 \%$ | $0 \%$ | $1 \%$ |
| 14 | $0 \%$ |  |  | $1 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $1 \%$ |
| $15+$ | $0 \%$ |  |  |  |  |  |  |  |  |


| AGE | 4.c | 5.a | 5.b | $\mathbf{6 . a}$ | $\mathbf{6 . b}$ | $\mathbf{7 . a}$ | $\mathbf{7 . b}$ | 7.c | 7.d |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | $1 \%$ |  | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $35 \%$ | $2 \%$ |
| 2 | $30 \%$ | $0 \%$ | $0 \%$ | $1 \%$ | $24 \%$ | $3 \%$ | $1 \%$ | $11 \%$ | $29 \%$ |
| 3 | $25 \%$ | $4 \%$ | $11 \%$ | $8 \%$ | $54 \%$ | $9 \%$ | $7 \%$ | $6 \%$ | $29 \%$ |
| 4 | $6 \%$ | $10 \%$ | $10 \%$ | $9 \%$ | $20 \%$ | $7 \%$ | $4 \%$ | $2 \%$ | $8 \%$ |
| 5 | $12 \%$ | $20 \%$ | $14 \%$ | $14 \%$ | $0 \%$ | $14 \%$ | $15 \%$ | $5 \%$ | $12 \%$ |
| 6 | $8 \%$ | $25 \%$ | $19 \%$ | $17 \%$ | $1 \%$ | $13 \%$ | $18 \%$ | $9 \%$ | $8 \%$ |
| 7 | $6 \%$ | $16 \%$ | $17 \%$ | $15 \%$ | $0 \%$ | $18 \%$ | $13 \%$ | $11 \%$ | $4 \%$ |
| 8 | $4 \%$ | $10 \%$ | $12 \%$ | $12 \%$ | $0 \%$ | $8 \%$ | $16 \%$ | $6 \%$ | $4 \%$ |
| 9 | $3 \%$ | $8 \%$ | $7 \%$ | $10 \%$ | $0 \%$ | $11 \%$ | $14 \%$ | $8 \%$ | $2 \%$ |
| 10 | $3 \%$ | $4 \%$ | $6 \%$ | $7 \%$ | $0 \%$ | $10 \%$ | $7 \%$ | $4 \%$ | $2 \%$ |
| 11 | $1 \%$ | $1 \%$ | $1 \%$ | $4 \%$ | $0 \%$ | $5 \%$ | $4 \%$ | $2 \%$ | $1 \%$ |
| 12 | $0 \%$ | $1 \%$ | $1 \%$ | $2 \%$ | $0 \%$ | $1 \%$ | $1 \%$ | $0 \%$ | $0 \%$ |
| 13 | $0 \%$ | $0 \%$ | $1 \%$ | $1 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ |
| 14 | $0 \%$ |  | $1 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ |
| $15+$ | $0 \%$ |  |  | $0 \%$ |  | $0 \%$ |  |  | $0 \%$ |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<1 \%$ (cont.).

Quarters 1-4

| Age | $7 . \mathrm{e}$ | 7.f | 7.9 | 7.h | 7.j | 7.k | 8.a | 8.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  | 55\% | 45\% |
| 1 | 8\% | 1\% | 70\% | 46\% | 7\% | 0\% | 11\% | 16\% |
| 2 | 11\% | 8\% | 17\% | 14\% | 2\% | 20\% | 10\% | 9\% |
| 3 | 27\% | 41\% | 4\% | 7\% | 4\% | 30\% | 7\% | 8\% |
| 4 | 12\% | 28\% | 1\% | 2\% | 4\% | 9\% | 3\% | 3\% |
| 5 | 8\% | 10\% | 1\% | 6\% | 10\% | 11\% | 4\% | 6\% |
| 6 | 6\% | 5\% | 2\% | 6\% | 18\% | 9\% | 4\% | 6\% |
| 7 | 5\% | 3\% | 2\% | 3\% | 18\% | 9\% | 3\% | 4\% |
| 8 | 5\% | 2\% | 1\% | 5\% | 13\% | 2\% | 2\% | 2\% |
| 9 | 4\% | 1\% | 1\% | 4\% | 12\% | 5\% | 1\% | 1\% |
| 10 | 4\% | 1\% | 0\% | 2\% | 5\% | 2\% | 1\% | 0\% |
| 11 | 5\% | 1\% | 0\% | 2\% | 4\% | 2\% | 0\% | 0\% |
| 12 | 2\% | 0\% | 0\% | 0\% | 2\% | 0\% | 0\% | 0\% |
| 13 | 2\% | 0\% | 0\% | 1\% | 1\% |  |  |  |
| 14 | 1\% | 0\% |  | 0\% | 1\% |  |  |  |
| 15+ |  |  |  |  |  |  |  |  |


| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | $\mathbf{1 4 . a}$ | $\mathbf{1 4 . b}$ | All |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | $0 \%$ |  | $17 \%$ | $12 \%$ | $5 \%$ |  |  | $1 \%$ |
| 1 | $4 \%$ | $3 \%$ | $14 \%$ | $36 \%$ | $25 \%$ | $3 \%$ | $3 \%$ | $1 \%$ |
| 2 | $7 \%$ | $5 \%$ | $5 \%$ | $10 \%$ | $19 \%$ | $10 \%$ | $9 \%$ | $3 \%$ |
| 3 | $12 \%$ | $6 \%$ | $6 \%$ | $22 \%$ | $19 \%$ | $44 \%$ | $42 \%$ | $15 \%$ |
| 4 | $8 \%$ | $7 \%$ | $3 \%$ | $9 \%$ | $8 \%$ | $28 \%$ | $27 \%$ | $12 \%$ |
| 5 | $17 \%$ | $19 \%$ | $13 \%$ | $3 \%$ | $7 \%$ | $12 \%$ | $13 \%$ | $13 \%$ |
| 6 | $19 \%$ | $22 \%$ | $11 \%$ | $2 \%$ | $6 \%$ | $2 \%$ | $3 \%$ | $17 \%$ |
| 7 | $15 \%$ | $17 \%$ | $12 \%$ | $2 \%$ | $4 \%$ | $0 \%$ | $1 \%$ | $14 \%$ |
| 8 | $9 \%$ | $10 \%$ | $4 \%$ | $1 \%$ | $3 \%$ |  | $1 \%$ | $9 \%$ |
| 9 | $6 \%$ | $6 \%$ | $6 \%$ | $1 \%$ | $2 \%$ |  | $1 \%$ | $6 \%$ |
| 10 | $2 \%$ | $2 \%$ | $4 \%$ | $1 \%$ | $1 \%$ |  | $0 \%$ | $5 \%$ |
| 11 | $1 \%$ | $1 \%$ | $3 \%$ | $0 \%$ | $0 \%$ |  | $0 \%$ | $2 \%$ |
| 12 | $0 \%$ | $0 \%$ | $0 \%$ | $1 \%$ | $0 \%$ |  |  | $1 \%$ |
| 13 | $0 \%$ | $0 \%$ |  |  | $0 \%$ |  |  | $0 \%$ |
| 14 | $0 \%$ | $0 \%$ |  |  |  |  |  | $0 \%$ |
| $15+$ |  | $0 \%$ |  |  |  |  |  | $0 \%$ |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<\mathbf{1 \%}$ (cont.).

Quarter 1

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | $3 . \mathrm{b}$ | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 0\% |  |  |  |  |  |  |  |  |
| 2 | 0\% |  |  |  |  |  | 0\% | 0\% | 0\% |
| 3 | 7\% |  |  | 16\% |  |  | 16\% | 16\% | 16\% |
| 4 | 7\% |  |  | 10\% |  |  | 11\% | 10\% | 11\% |
| 5 | 15\% |  |  | 8\% |  |  | 8\% | 9\% | 13\% |
| 6 | 22\% |  |  | 14\% |  |  | 13\% | 14\% | 16\% |
| 7 | 18\% |  |  | 18\% |  |  | 18\% | 17\% | 16\% |
| 8 | 11\% |  |  | 14\% |  |  | 13\% | 14\% | 13\% |
| 9 | 8\% |  |  | 6\% |  |  | 5\% | 6\% | 5\% |
| 10 | 6\% |  |  | 9\% |  |  | 8\% | 8\% | 5\% |
| 11 | 2\% |  |  | 0\% |  |  | 0\% | 0\% | 0\% |
| 12 | 2\% |  |  | 2\% |  |  | 3\% | 2\% | 3\% |
| 13 | 1\% |  |  | 2\% |  |  | 3\% | 2\% | 3\% |
| 14 | 0\% |  |  | 2\% |  |  | 3\% | 2\% |  |
| 15+ | 0\% |  |  | 0\% |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| AGE | 4.c | 5.a | 5.b | $6 . a$ | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  | 0\% |  | 0\% | 0\% | 40\% | 3\% |
| 2 | 2\% | 1\% |  | 1\% |  | 0\% | 1\% | 10\% | 2\% |
| 3 | 18\% | 5\% | 16\% | 8\% |  | 4\% | 7\% | 2\% | 18\% |
| 4 | 11\% | 11\% | 10\% | 10\% |  | 6\% | 4\% | 2\% | 10\% |
| 5 | 18\% | 22\% | 8\% | 14\% |  | 11\% | 15\% | 4\% | 17\% |
| 6 | 16\% | 26\% | 14\% | 17\% |  | 15\% | 18\% | 9\% | 16\% |
| 7 | 14\% | 16\% | 18\% | 15\% |  | 19\% | 13\% | 12\% | 14\% |
| 8 | 9\% | 9\% | 14\% | 12\% |  | 13\% | 16\% | 6\% | 9\% |
| 9 | 7\% | 5\% | 6\% | 10\% |  | 11\% | 14\% | 9\% | 7\% |
| 10 | 3\% | 3\% | 9\% | 7\% |  | 12\% | 7\% | 4\% | 3\% |
| 11 | 0\% | 1\% | 0\% | 4\% |  | 5\% | 4\% | 3\% | 0\% |
| 12 | 1\% | 0\% | 2\% | 2\% |  | 3\% | 1\% |  | 1\% |
| 13 | 0\% |  | 2\% | 1\% |  | 1\% | 0\% |  | 0\% |
| 14 |  |  | 2\% | 0\% |  | 0\% |  |  |  |
| 15+ |  |  |  | 0\% |  | 0\% |  |  |  |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<1 \%$ (cont.).

## Quarter 1

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  | $43 \%$ | $22 \%$ |
| 1 | $24 \%$ | $1 \%$ | $80 \%$ | $56 \%$ | $8 \%$ |  | $18 \%$ | $21 \%$ |
| 2 | $8 \%$ | $4 \%$ | $18 \%$ | $13 \%$ | $2 \%$ |  | $9 \%$ | $13 \%$ |
| 3 | $17 \%$ | $46 \%$ | $1 \%$ | $4 \%$ | $4 \%$ |  | $8 \%$ | $12 \%$ |
| 4 | $11 \%$ | $33 \%$ | $0 \%$ | $1 \%$ | $4 \%$ |  | $3 \%$ | $5 \%$ |
| 5 | $10 \%$ | $10 \%$ | $0 \%$ | $5 \%$ | $10 \%$ |  | $6 \%$ | $9 \%$ |
| 6 | $9 \%$ | $4 \%$ | $0 \%$ | $5 \%$ | $19 \%$ | $20 \%$ | $6 \%$ | $9 \%$ |
| 7 | $5 \%$ | $2 \%$ | $0 \%$ | $3 \%$ | $18 \%$ | $30 \%$ | $4 \%$ | $5 \%$ |
| 8 | $7 \%$ | $1 \%$ | $0 \%$ | $5 \%$ | $13 \%$ | $10 \%$ | $2 \%$ | $3 \%$ |
| 9 | $5 \%$ | $1 \%$ | $0 \%$ | $4 \%$ | $13 \%$ | $20 \%$ | $1 \%$ | $1 \%$ |
| 10 | $3 \%$ | $0 \%$ | $0 \%$ | $2 \%$ | $6 \%$ | $10 \%$ | $0 \%$ | $0 \%$ |
| 11 | $2 \%$ |  |  | $1 \%$ | $2 \%$ | $10 \%$ | $0 \%$ | $0 \%$ |
| 12 | $0 \%$ |  |  |  |  |  |  |  |
| 13 |  |  |  |  |  |  |  | $0 \%$ |
| 14 |  |  |  |  |  |  |  |  |
| $15+$ |  |  |  |  |  |  |  |  |


| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b | All |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  | $3 \%$ |  |  | $1 \%$ |  |  |
| 1 | $4 \%$ | $4 \%$ | $16 \%$ | $11 \%$ | $0 \%$ | $1 \%$ |  |  |
| 2 | $9 \%$ | $6 \%$ | $8 \%$ | $14 \%$ | $3 \%$ | $2 \%$ |  |  |
| 3 | $13 \%$ | $6 \%$ | $11 \%$ | $25 \%$ | $21 \%$ | $9 \%$ |  |  |
| 4 | $8 \%$ | $7 \%$ | $6 \%$ | $22 \%$ | $15 \%$ | $9 \%$ |  |  |
| 5 | $18 \%$ | $19 \%$ | $14 \%$ | $2 \%$ | $25 \%$ | $14 \%$ |  |  |
| 6 | $19 \%$ | $22 \%$ | $16 \%$ | $2 \%$ | $21 \%$ | $17 \%$ |  |  |
| 7 | $14 \%$ | $17 \%$ | $12 \%$ | $6 \%$ | $9 \%$ | $15 \%$ |  |  |
| 8 | $8 \%$ | $10 \%$ | $7 \%$ | $4 \%$ | $3 \%$ | $11 \%$ |  |  |
| 9 | $5 \%$ | $6 \%$ | $4 \%$ | $3 \%$ | $2 \%$ | $9 \%$ |  |  |
| 10 | $2 \%$ | $2 \%$ | $2 \%$ | $7 \%$ | $1 \%$ | $7 \%$ |  |  |
| 11 | $0 \%$ | $1 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $3 \%$ |  |  |
| 12 | $0 \%$ | $0 \%$ | $0 \%$ | $3 \%$ | $0 \%$ | $1 \%$ |  |  |
| 13 |  | $0 \%$ |  |  |  | $2 \%$ |  |  |
| 14 |  |  |  |  |  | $0 \%$ |  |  |
| $15+$ |  |  |  |  |  |  |  |  |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<\mathbf{1 \%}$ (cont.).

Quarter 2

| Age | 2.a | 2.a.1 | 2.a.2 | 3.a | 3.b | 3.c | 3.d | 4.a |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  | 4.b |  |
| 1 | $0 \%$ | $0 \%$ |  |  |  | $0 \%$ | $0 \%$ |  |
| 2 | $0 \%$ | $0 \%$ |  |  | $2 \%$ | $15 \%$ |  |  |
| 3 | $28 \%$ | $13 \%$ | $16 \%$ | $6 \%$ | $15 \%$ | $11 \%$ | $34 \%$ |  |
| 4 | $21 \%$ | $8 \%$ | $10 \%$ | $6 \%$ | $9 \%$ | $3 \%$ | $7 \%$ |  |
| 5 | $12 \%$ | $8 \%$ | $8 \%$ | $13 \%$ | $8 \%$ | $9 \%$ | $7 \%$ |  |
| 6 | $13 \%$ | $15 \%$ | $14 \%$ | $19 \%$ | $14 \%$ | $17 \%$ | $9 \%$ |  |
| 7 | $16 \%$ | $18 \%$ | $17 \%$ | $19 \%$ | $18 \%$ | $19 \%$ | $9 \%$ |  |
| 8 | $4 \%$ | $13 \%$ | $14 \%$ | $6 \%$ | $13 \%$ | $13 \%$ | $8 \%$ |  |
| 9 | $2 \%$ | $7 \%$ | $6 \%$ | $13 \%$ | $6 \%$ | $10 \%$ | $3 \%$ |  |
| 10 | $1 \%$ | $9 \%$ | $9 \%$ | $6 \%$ | $8 \%$ | $8 \%$ | $5 \%$ |  |
| 11 | $1 \%$ | $2 \%$ | $0 \%$ | $13 \%$ | $1 \%$ | $4 \%$ | $0 \%$ |  |
| 12 | $1 \%$ | $2 \%$ | $2 \%$ |  | $2 \%$ | $2 \%$ | $1 \%$ |  |
| 13 |  | $2 \%$ | $2 \%$ |  | $2 \%$ | $1 \%$ | $1 \%$ |  |
| 14 |  | $2 \%$ | $2 \%$ |  | $2 \%$ | $0 \%$ | $1 \%$ |  |
| $15+$ |  | $0 \%$ | $0 \%$ |  |  | $0 \%$ |  |  |


| AGE | 4.c | 5.a | 5.b | $\mathbf{6 . a}$ | $\mathbf{6 . b}$ | 7.a | 7.b | 7.c | 7.d |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | $1 \%$ |  |  |  |  | $0 \%$ |  |  | $2 \%$ |
| 2 | $26 \%$ | $1 \%$ | $2 \%$ | $3 \%$ | $27 \%$ | $6 \%$ | $1 \%$ | $27 \%$ | $35 \%$ |
| 3 | $23 \%$ | $5 \%$ | $15 \%$ | $3 \%$ | $37 \%$ | $29 \%$ | $8 \%$ | $36 \%$ | $29 \%$ |
| 4 | $6 \%$ | $11 \%$ | $10 \%$ | $2 \%$ | $10 \%$ | $18 \%$ | $4 \%$ | $10 \%$ | $7 \%$ |
| 5 | $13 \%$ | $22 \%$ | $19 \%$ | $18 \%$ | $13 \%$ | $12 \%$ | $16 \%$ | $13 \%$ | $11 \%$ |
| 6 | $8 \%$ | $26 \%$ | $24 \%$ | $14 \%$ | $9 \%$ | $9 \%$ | $17 \%$ | $9 \%$ | $7 \%$ |
| 7 | $8 \%$ | $16 \%$ | $13 \%$ | $22 \%$ | $3 \%$ | $6 \%$ | $11 \%$ | $3 \%$ | $3 \%$ |
| 8 | $4 \%$ | $9 \%$ | $7 \%$ | $6 \%$ | $1 \%$ | $6 \%$ | $16 \%$ | $1 \%$ | $3 \%$ |
| 9 | $5 \%$ | $5 \%$ | $5 \%$ | $14 \%$ | $0 \%$ | $6 \%$ | $13 \%$ | $1 \%$ | $2 \%$ |
| 10 | $4 \%$ | $3 \%$ | $3 \%$ | $11 \%$ | $0 \%$ | $7 \%$ | $7 \%$ | $0 \%$ | $2 \%$ |
| 11 | $2 \%$ | $1 \%$ | $1 \%$ | $7 \%$ |  | $1 \%$ | $5 \%$ | $0 \%$ | $1 \%$ |
| 12 | $0 \%$ | $0 \%$ | $1 \%$ | $0 \%$ |  |  | $1 \%$ |  | $0 \%$ |
| 13 | $0 \%$ |  | $0 \%$ | $0 \%$ |  |  | $1 \%$ |  |  |
| 14 | $0 \%$ |  |  | $0 \%$ |  |  | $0 \%$ |  |  |
| $15+$ | $0 \%$ |  |  |  |  |  | $0 \%$ |  | $0 \%$ |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<1 \%$ (cont.).

Quarter 2

| Age | 7.e | 7.f | $7 . \mathrm{g}$ | 7.h | 7.j | 7.k | 8.a | 8.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  | 35\% | 50\% |
| 1 |  | 0\% |  |  |  |  | 15\% | 21\% |
| 2 | 12\% | 6\% | 7\% | 20\% | 1\% | 28\% | 2\% | 2\% |
| 3 | 29\% | 32\% | 13\% | 29\% | 4\% | 38\% | 3\% | 2\% |
| 4 | 12\% | 20\% | 6\% | 8\% | 2\% | 9\% | 3\% | 2\% |
| 5 | 13\% | 11\% | 12\% | 14\% | 5\% | 13\% | 10\% | 5\% |
| 6 | 11\% | 9\% | 17\% | 11\% | 8\% | 9\% | 11\% | 6\% |
| 7 | 5\% | 4\% | 15\% | 5\% | 13\% | 3\% | 10\% | 5\% |
| 8 | 7\% | 6\% | 12\% | 5\% | 13\% |  | 6\% | 3\% |
| 9 | 6\% | 5\% | 11\% | 4\% | 8\% |  | 4\% | 2\% |
| 10 | 3\% | 7\% | 5\% | 2\% | 5\% |  | 1\% | 1\% |
| 11 | 2\% | 0\% | 2\% | 1\% | 20\% |  | 0\% | 0\% |
| 12 | 0\% |  | 1\% | 0\% | 7\% |  |  |  |
| 13 |  |  | 0\% |  | 9\% |  |  |  |
| 14 |  |  | 0\% |  | 5\% |  |  |  |
| 15+ |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |
| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b | All |
| 0 |  |  | 30\% |  |  |  |  | 4\% |
| 1 | 3\% | 0\% | 12\% | 56\% | 4\% |  | 3\% | 3\% |
| 2 | 4\% | 1\% | 3\% | 3\% | 23\% |  | 10\% | 4\% |
| 3 | 12\% | 4\% | 1\% | 22\% | 22\% |  | 44\% | 19\% |
| 4 | 7\% | 7\% | 0\% | 7\% | 6\% |  | 28\% | 12\% |
| 5 | 16\% | 20\% | 12\% | 4\% | 9\% |  | 12\% | 13\% |
| 6 | 19\% | 23\% | 7\% | 3\% | 11\% |  | 2\% | 14\% |
| 7 | 16\% | 19\% | 13\% | 2\% | 10\% |  | 0\% | 14\% |
| 8 | 11\% | 11\% | 2\% | 2\% | 7\% |  | 0\% | 7\% |
| 9 | 7\% | 8\% | 9\% | 1\% | 5\% |  |  | 5\% |
| 10 | 3\% | 3\% | 6\% | 1\% | 2\% |  |  | 3\% |
| 11 | 1\% | 2\% | 4\% | 0\% | 1\% |  |  | 1\% |
| 12 | 0\% | 0\% |  |  | 0\% |  |  | 1\% |
| 13 | 0\% | 1\% |  |  | 0\% |  |  | 0\% |
| 14 |  |  |  |  |  |  |  | 0\% |
| 15+ |  |  |  |  |  |  |  | 0\% |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<\mathbf{1 \%}$ (cont.).

Quarter 3

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | 3.b | 3.c | 3.d | $4 . \mathrm{a}$ | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 0\% |  |  | 2\% | 4\% | 2\% | 2\% | 4\% | 1\% |
| 2 | 0\% |  |  | 4\% | 6\% | 2\% | 4\% | 6\% | 6\% |
| 3 | 16\% | 4\% | 4\% | 17\% | 24\% | 19\% | 18\% | 24\% | 58\% |
| 4 | 13\% | 6\% | 6\% | 12\% | 1\% | 8\% | 22\% | 8\% | 14\% |
| 5 | 12\% | 12\% | 12\% | 15\% | 7\% | 15\% | 17\% | 9\% | 9\% |
| 6 | 19\% | 34\% | 34\% | 16\% | 18\% | 21\% | 17\% | 13\% | 7\% |
| 7 | 18\% | 24\% | 24\% | 13\% | 18\% | 14\% | 10\% | 12\% | 3\% |
| 8 | 9\% | 12\% | 12\% | 9\% | 10\% | 10\% | 4\% | 8\% | 1\% |
| 9 | 6\% | 6\% | 6\% | 5\% | 6\% | 5\% | 3\% | 7\% | 1\% |
| 10 | 4\% | 2\% | 2\% | 4\% | 6\% | 3\% | 2\% | 6\% | 0\% |
| 11 | 2\% |  |  | 1\% | 2\% | 1\% | 1\% | 2\% | 0\% |
| 12 | 1\% |  |  | 1\% |  | 0\% |  | 1\% | 0\% |
| 13 | 0\% |  |  | 0\% |  | 0\% |  | 0\% |  |
| 14 | 0\% |  |  | 0\% |  | 0\% |  | 0\% |  |
| 15+ | 0\% |  |  |  |  |  |  | 0\% |  |
| AGE | 4.c | 5.a | 5.b | 6.9 | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 2\% |  |  | 0\% |  |  | 0\% |  | 2\% |
| 2 | 35\% | 0\% | 0\% | 4\% | 24\% | 5\% | 0\% | 19\% | 35\% |
| 3 | 29\% | 4\% | 4\% | 1\% | 54\% | 8\% | 1\% | 45\% | 30\% |
| 4 | 7\% | 10\% | 10\% | 1\% | 20\% | 7\% | 1\% | 16\% | 7\% |
| 5 | 11\% | 20\% | 20\% | 20\% | 0\% | 18\% | 1\% |  | 11\% |
| 6 | 7\% | 25\% | 25\% | 12\% | 0\% | 11\% | 1\% |  | 7\% |
| 7 | 3\% | 16\% | 16\% | 23\% | 0\% | 19\% | 12\% | 3\% | 2\% |
| 8 | 3\% | 10\% | 10\% | 4\% | 0\% | 4\% | 12\% | 3\% | 3\% |
| 9 | 2\% | 8\% | 8\% | 15\% | 0\% | 13\% | 4\% | 0\% | 1\% |
| 10 | 1\% | 4\% | 4\% | 11\% | 0\% | 9\% | 4\% | 0\% | 1\% |
| 11 | 1\% | 2\% | 2\% | 8\% |  | 6\% | 30\% | 6\% | 1\% |
| 12 | 0\% | 1\% | 1\% |  |  |  | 11\% | 3\% | 0\% |
| 13 |  |  |  |  |  |  | 15\% | 3\% | 0\% |
| 14 |  |  |  |  |  |  | 7\% |  | 0\% |
| 15+ | 0\% |  |  |  |  |  |  |  | 0\% |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<1 \%$ (cont.).

Quarter 3

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  | $72 \%$ | $87 \%$ |
| 1 | $1 \%$ | $0 \%$ |  |  |  |  | $1 \%$ | $1 \%$ |
| 2 | $11 \%$ | $7 \%$ | $21 \%$ | $2 \%$ | $20 \%$ |  | $7 \%$ | $4 \%$ |
| 3 | $28 \%$ | $35 \%$ | $49 \%$ | $3 \%$ | $46 \%$ | $24 \%$ | $10 \%$ | $4 \%$ |
| 4 | $14 \%$ | $29 \%$ | $19 \%$ | $1 \%$ | $17 \%$ | $55 \%$ | $5 \%$ | $2 \%$ |
| 5 | $6 \%$ | $12 \%$ | $1 \%$ | $1 \%$ | $0 \%$ | $20 \%$ | $1 \%$ | $1 \%$ |
| 6 | $4 \%$ | $6 \%$ | $1 \%$ | $1 \%$ | $0 \%$ |  | $1 \%$ | $0 \%$ |
| 7 | $6 \%$ | $3 \%$ | $2 \%$ | $11 \%$ | $2 \%$ |  | $1 \%$ | $0 \%$ |
| 8 | $5 \%$ | $2 \%$ | $2 \%$ | $11 \%$ | $2 \%$ |  | $0 \%$ | $0 \%$ |
| 9 | $2 \%$ | $2 \%$ | $0 \%$ | $4 \%$ | $1 \%$ |  |  |  |
| 10 | $2 \%$ | $1 \%$ | $0 \%$ | $4 \%$ | $1 \%$ |  | $1 \%$ | $0 \%$ |
| 11 | $10 \%$ | $1 \%$ | $2 \%$ | $29 \%$ | $5 \%$ |  |  |  |
| 12 | $4 \%$ | $0 \%$ | $1 \%$ | $11 \%$ | $2 \%$ |  | $0 \%$ | $0 \%$ |
| 13 | $5 \%$ | $0 \%$ | $1 \%$ | $14 \%$ | $3 \%$ |  |  |  |
| 14 | $3 \%$ | $0 \%$ | $0 \%$ | $7 \%$ | $1 \%$ |  |  |  |
| $15+$ |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |


| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b | All |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | $5 \%$ |  |  | $29 \%$ | $25 \%$ |  |  | $1 \%$ |
| 1 | $14 \%$ | $0 \%$ | $8 \%$ | $27 \%$ | $16 \%$ | $3 \%$ | $3 \%$ | $0 \%$ |
| 2 | $19 \%$ | $22 \%$ | $8 \%$ | $15 \%$ | $15 \%$ | $10 \%$ | $9 \%$ | $1 \%$ |
| 3 | $40 \%$ | $44 \%$ | $46 \%$ | $22 \%$ | $25 \%$ | $44 \%$ | $42 \%$ | $16 \%$ |
| 4 | $15 \%$ | $23 \%$ | $15 \%$ | $5 \%$ | $13 \%$ | $28 \%$ | $27 \%$ | $13 \%$ |
| 5 | $4 \%$ | $6 \%$ | $8 \%$ | $2 \%$ | $4 \%$ | $12 \%$ | $13 \%$ | $13 \%$ |
| 6 | $2 \%$ | $3 \%$ | $8 \%$ |  | $1 \%$ | $2 \%$ | $3 \%$ | $19 \%$ |
| 7 | $1 \%$ | $1 \%$ | $8 \%$ |  | $0 \%$ |  | $1 \%$ | $16 \%$ |
| 8 | $1 \%$ | $1 \%$ |  |  | $0 \%$ |  | $1 \%$ | $8 \%$ |
| 9 | $0 \%$ | $0 \%$ |  |  |  |  | $1 \%$ | $6 \%$ |
| 10 | $0 \%$ | $0 \%$ |  |  |  |  | $0 \%$ | $4 \%$ |
| 11 |  |  |  |  |  |  |  | $0 \%$ |
| 12 | $0 \%$ |  |  |  |  |  |  |  |
| 13 |  |  |  |  |  |  |  |  |
| 14 |  |  |  |  |  |  |  | $1 \%$ |
| $15+$ |  |  |  |  |  |  |  | $0 \%$ |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<\mathbf{1 \%}$ (cont.).

Quarter 4

| Age | 2.a | 2.a.1 | 2.a.2 | 3.a | 3.b | 3.c | 3.d | 4.a | 4.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | $3 \%$ | $2 \%$ | $2 \%$ | $2 \%$ | $5 \%$ | $2 \%$ | $1 \%$ | $2 \%$ | $3 \%$ |
| 2 | $1 \%$ | $4 \%$ | $4 \%$ | $4 \%$ | $5 \%$ | $2 \%$ | $7 \%$ | $7 \%$ | $12 \%$ |
| 3 | $17 \%$ | $17 \%$ | $17 \%$ | $17 \%$ | $24 \%$ | $19 \%$ | $15 \%$ | $21 \%$ | $41 \%$ |
| 4 | $6 \%$ | $17 \%$ | $17 \%$ | $14 \%$ | $5 \%$ | $8 \%$ | $21 \%$ | $14 \%$ | $15 \%$ |
| 5 | $9 \%$ | $18 \%$ | $18 \%$ | $16 \%$ | $10 \%$ | $15 \%$ | $17 \%$ | $14 \%$ | $12 \%$ |
| 6 | $16 \%$ | $14 \%$ | $14 \%$ | $17 \%$ | $19 \%$ | $21 \%$ | $14 \%$ | $14 \%$ | $8 \%$ |
| 7 | $16 \%$ | $9 \%$ | $9 \%$ | $13 \%$ | $14 \%$ | $14 \%$ | $12 \%$ | $11 \%$ | $4 \%$ |
| 8 | $11 \%$ | $2 \%$ | $2 \%$ | $8 \%$ | $10 \%$ | $10 \%$ | $7 \%$ | $7 \%$ | $2 \%$ |
| 9 | $9 \%$ | $6 \%$ | $6 \%$ | $4 \%$ | $5 \%$ | $5 \%$ | $3 \%$ | $4 \%$ | $1 \%$ |
| 10 | $8 \%$ | $7 \%$ | $7 \%$ | $3 \%$ | $5 \%$ | $4 \%$ | $2 \%$ | $3 \%$ | $1 \%$ |
| 11 | $3 \%$ | $2 \%$ | $2 \%$ | $1 \%$ |  | $1 \%$ | $1 \%$ | $1 \%$ | $0 \%$ |
| 12 | $2 \%$ | $2 \%$ | $2 \%$ | $0 \%$ |  | $0 \%$ |  | $1 \%$ | $0 \%$ |
| 13 | $0 \%$ |  |  | $0 \%$ |  |  |  | $0 \%$ | $0 \%$ |
| 14 | $0 \%$ |  |  | $0 \%$ |  | $0 \%$ |  | $0 \%$ |  |
| $15+$ | $0 \%$ |  |  |  |  |  |  |  |  |


| AGE | 4.c | 5.a | 5.b | 6.9 | 6.b | 7.a | 7.b | 7.c | 7.d |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  | 0\% |  |  |  |
| 1 |  |  | 4\% | 2\% |  | 2\% |  |  | 2\% |
| 2 | 40\% | 1\% | 6\% | 4\% |  | 8\% | 0\% |  | 32\% |
| 3 | 26\% | 19\% | 25\% | 15\% |  | 29\% | 0\% |  | 32\% |
| 4 | 6\% | 10\% |  | 2\% |  | 16\% | 0\% |  | 7\% |
| 5 | 11\% | 19\% | 6\% | 5\% |  | 13\% | 0\% |  | 11\% |
| 6 | 7\% | 25\% | 18\% | 11\% |  | 13\% | 0\% |  | 7\% |
| 7 | 2\% | 12\% | 18\% | 15\% |  | 9\% | 12\% |  | 3\% |
| 8 | 3\% | 6\% | 10\% | 10\% |  | 6\% | 12\% |  | 2\% |
| 9 | 2\% | 4\% | 6\% | 5\% |  | 3\% | 4\% |  | 2\% |
| 10 | 1\% | 3\% | 6\% | 5\% |  | 2\% | 4\% |  | 1\% |
| 11 | 1\% | 1\% | 2\% | 12\% |  |  | 31\% |  | 1\% |
| 12 | 0\% | 1\% | 0\% | 4\% |  |  | 12\% |  | 0\% |
| 13 |  | 0\% |  | 6\% |  |  | 16\% |  |  |
| 14 |  |  |  | 3\% |  |  | 8\% |  |  |
| 15+ | 0\% |  |  |  |  |  |  |  | 0\% |

Table 8.4.3.2. NE Atlantic Mackerel. Percentage catch numbers-at-age by area for 2017. Zeros represent values $<1 \%$ (cont.).

## Quarter 4

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  | $56 \%$ |
| 1 | $2 \%$ | $2 \%$ | $2 \%$ |  |  | $83 \%$ |  |
| 2 | $13 \%$ | $13 \%$ | $2 \%$ | $1 \%$ | $12 \%$ | $16 \%$ | $4 \%$ |
| 3 | $35 \%$ | $37 \%$ | $17 \%$ | $3 \%$ | $29 \%$ | $19 \%$ | $7 \%$ |
| 4 | $12 \%$ | $22 \%$ | $8 \%$ | $1 \%$ | $10 \%$ | $6 \%$ | $4 \%$ |
| 5 | $6 \%$ | $10 \%$ | $14 \%$ |  |  | $2 \%$ | $2 \%$ |
| 6 | $5 \%$ | $5 \%$ | $19 \%$ |  |  | $0 \%$ | $0 \%$ |
| 7 | $4 \%$ | $3 \%$ | $14 \%$ | $12 \%$ | $6 \%$ |  | $0 \%$ |
| 8 | $4 \%$ | $2 \%$ | $10 \%$ | $12 \%$ | $6 \%$ | $0 \%$ | $0 \%$ |
| 9 | $4 \%$ | $1 \%$ | $5 \%$ | $4 \%$ | $2 \%$ |  |  |
| 10 | $5 \%$ | $1 \%$ | $3 \%$ | $4 \%$ | $2 \%$ |  |  |
| 11 | $5 \%$ | $1 \%$ | $3 \%$ | $31 \%$ | $14 \%$ |  |  |
| 12 | $2 \%$ | $0 \%$ | $2 \%$ | $12 \%$ | $6 \%$ |  |  |
| 13 | $3 \%$ | $1 \%$ | $2 \%$ | $15 \%$ | $8 \%$ |  |  |
| 14 | $1 \%$ | $0 \%$ | $0 \%$ | $7 \%$ | $4 \%$ |  |  |
| $15+$ |  |  |  |  |  |  |  |


| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | $55 \%$ |  |  | $29 \%$ | $2 \%$ |  | All |
| 1 | $14 \%$ | $3 \%$ |  | $24 \%$ | $55 \%$ | $1 \%$ |  |
| 2 | $14 \%$ | $32 \%$ | $14 \%$ | $22 \%$ | $3 \%$ |  |  |
| 3 | $13 \%$ | $42 \%$ | $22 \%$ | $14 \%$ | $6 \%$ |  |  |
| 4 | $3 \%$ | $16 \%$ | $10 \%$ | $6 \%$ | $20 \%$ |  |  |
| 5 | $0 \%$ | $3 \%$ | $1 \%$ | $1 \%$ | $12 \%$ |  |  |
| 6 | $0 \%$ | $2 \%$ | $1 \%$ | $0 \%$ | $13 \%$ |  |  |
| 7 | $0 \%$ | $1 \%$ | $0 \%$ | $0 \%$ | $14 \%$ |  |  |
| 8 |  | $1 \%$ |  |  | $12 \%$ |  |  |
| 9 |  |  |  |  |  | $8 \%$ |  |
| 10 |  | $2 \%$ |  |  | $5 \%$ |  |  |
| 11 |  |  |  |  |  | $4 \%$ |  |
| 12 |  |  |  |  |  | $2 \%$ |  |
| 13 |  |  |  |  |  |  |  |
| 14 |  |  |  |  | $0 \%$ |  |  |
| $15+$ |  |  |  |  | $0 \%$ |  |  |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017.

## Quarters 1-4

| Age | $2 . a$ | 2.a. 1 | 2.a. 2 | 3.a | 3.b | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 291 | 335 | 335 | 284 | 280 | 282 | 292 | 288 | 282 |
| 2 | 328 | 335 | 335 | 323 | 311 | 321 | 320 | 317 | 313 |
| 3 | 324 | 345 | 331 | 326 | 327 | 334 | 328 | 328 | 316 |
| 4 | 328 | 355 | 355 | 346 | 344 | 352 | 350 | 348 | 341 |
| 5 | 345 | 362 | 366 | 351 | 347 | 357 | 356 | 353 | 344 |
| 6 | 353 | 370 | 372 | 357 | 346 | 360 | 364 | 360 | 351 |
| 7 | 357 | 386 | 376 | 361 | 353 | 368 | 367 | 365 | 354 |
| 8 | 367 | 396 | 379 | 373 | 365 | 378 | 383 | 374 | 368 |
| 9 | 374 | 391 | 386 | 375 | 370 | 381 | 375 | 377 | 368 |
| 10 | 379 | 403 | 395 | 383 | 381 | 385 | 384 | 380 | 385 |
| 11 | 384 | 395 | 395 | 383 | 390 | 389 | 385 | 386 | 393 |
| 12 | 389 | 395 | 395 | 386 | 383 | 393 | 382 | 390 | 382 |
| 13 | 393 |  |  | 385 | 382 | 410 | 381 | 392 | 381 |
| 14 | 398 |  |  | 366 | 360 | 386 | 362 | 371 | 361 |
| 15+ | 408 |  |  | 409 |  |  |  | 411 | 412 |
|  |  |  |  |  |  |  |  |  |  |
| AGE | 4.c | $5 . \mathrm{a}$ | 5.b | 6.9 | $6 . \mathrm{b}$ | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  | 165 |  | 165 |  |  |  |
| 1 | 271 |  | 280 | 274 |  | 242 | 298 | 298 | 278 |
| 2 | 311 | 338 | 327 | 291 | 335 | 291 | 304 | 341 | 311 |
| 3 | 317 | 338 | 313 | 313 | 335 | 297 | 316 | 356 | 317 |
| 4 | 334 | 348 | 342 | 340 | 335 | 317 | 321 | 371 | 335 |
| 5 | 342 | 354 | 350 | 348 | 357 | 345 | 344 | 362 | 343 |
| 6 | 348 | 358 | 355 | 353 | 360 | 350 | 346 | 375 | 351 |
| 7 | 359 | 364 | 357 | 360 | 366 | 359 | 362 | 383 | 365 |
| 8 | 365 | 373 | 370 | 369 | 374 | 368 | 368 | 381 | 369 |
| 9 | 377 | 376 | 372 | 379 | 376 | 382 | 373 | 392 | 376 |
| 10 | 384 | 382 | 385 | 382 | 382 | 388 | 379 | 394 | 378 |
| 11 | 396 | 387 | 387 | 385 | 387 | 389 | 371 | 405 | 397 |
| 12 | 393 | 393 | 383 | 391 | 393 | 391 | 392 | 404 | 399 |
| 13 | 388 | 420 | 380 | 393 | 420 | 390 | 407 | 410 | 395 |
| 14 | 373 |  | 360 | 406 |  | 385 | 416 | 419 | 406 |
| 15+ | 413 |  |  | 399 |  | 381 |  |  | 413 |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017 (cont.).

Quarters 1-4

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  | 137 | 136 |
| 1 | 295 | 231 | 298 | 298 | 298 | 0 | 215 | 190 |
| 2 | 299 | 268 | 342 | 343 | 338 | 348 | 304 | 303 |
| 3 | 310 | 291 | 349 | 349 | 320 | 367 | 334 | 328 |
| 4 | 318 | 305 | 347 | 346 | 336 | 377 | 349 | 346 |
| 5 | 339 | 317 | 357 | 353 | 347 | 384 | 357 | 354 |
| 6 | 347 | 325 | 356 | 351 | 352 | 388 | 364 | 359 |
| 7 | 368 | 334 | 364 | 370 | 362 | 390 | 373 | 369 |
| 8 | 371 | 352 | 375 | 369 | 375 | 390 | 382 | 379 |
| 9 | 380 | 351 | 377 | 372 | 377 | 396 | 389 | 386 |
| 10 | 388 | 386 | 382 | 380 | 383 | 397 | 398 | 390 |
| 11 | 401 | 405 | 384 | 385 | 391 | 409 | 402 | 400 |
| 12 | 408 | 408 | 391 | 408 | 394 | 445 | 427 | 417 |
| 13 | 410 | 410 | 408 | 410 | 409 |  | 438 | 437 |
| 14 | 420 | 420 | 416 | 420 | 418 |  |  |  |
| $15+$ | 413 |  |  |  |  |  |  |  |


| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b | All |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 181 |  | 131 | 189 | 212 |  |  | 138 |
| 1 | 257 | 263 | 185 | 219 | 256 | 352 | 352 | 269 |
| 2 | 296 | 299 | 299 | 313 | 292 | 336 | 336 | 315 |
| 3 | 321 | 327 | 319 | 333 | 324 | 357 | 357 | 327 |
| 4 | 345 | 347 | 345 | 364 | 343 | 372 | 371 | 340 |
| 5 | 355 | 354 | 352 | 371 | 351 | 383 | 381 | 350 |
| 6 | 361 | 361 | 357 | 371 | 357 | 400 | 380 | 355 |
| 7 | 371 | 372 | 363 | 390 | 372 | 422 | 367 | 360 |
| 8 | 380 | 379 | 377 | 389 | 382 |  | 374 | 370 |
| 9 | 385 | 386 | 385 | 394 | 388 |  | 376 | 376 |
| 10 | 389 | 390 | 394 | 400 | 393 |  | 382 | 381 |
| 11 | 403 | 407 | 395 | 408 | 403 |  | 387 | 385 |
| 12 | 414 | 416 | 405 | 393 | 412 |  | 394 | 390 |
| 13 | 433 | 436 | 418 |  | 440 |  |  | 393 |
| 14 |  |  | 372 |  |  |  | 394 |  |
| $15+$ |  |  |  |  |  |  |  | 306 |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017 (cont.).

## Quarter 1



Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017 (cont.).

Quarter 1

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  | 129 | 129 |
| 1 | 297 | 257 | 298 | 298 | 298 |  | 174 | 188 |
| 2 | 326 | 279 | 342 | 341 | 338 |  | 302 | 302 |
| 3 | 304 | 296 | 359 | 328 | 317 |  | 325 | 326 |
| 4 | 310 | 303 | 365 | 317 | 336 | 395 | 344 | 345 |
| 5 | 335 | 312 | 360 | 343 | 347 | 355 | 352 | 353 |
| 6 | 342 | 316 | 355 | 343 | 352 | 380 | 357 | 358 |
| 7 | 357 | 326 | 361 | 362 | 361 | 385 | 367 | 367 |
| 8 | 366 | 347 | 375 | 366 | 374 | 383 | 378 | 378 |
| 9 | 370 | 350 | 378 | 371 | 377 | 394 | 384 | 384 |
| 10 | 378 | 417 | 382 | 377 | 381 | 395 | 388 | 387 |
| 11 | 370 | 367 | 380 | 367 | 377 | 408 | 399 | 399 |
| 12 | 394 | 405 | 386 | 405 | 386 |  | 411 | 410 |
| 13 | 401 | 405 | 407 | 405 | 407 |  | 435 | 435 |
| 14 | 372 |  | 405 |  | 405 |  |  |  |
| $15+$ |  |  |  |  |  |  |  |  |

15+

| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  | 151 |  |  | All |  |
| 1 | 268 | 263 | 197 | 192 | 269 | 130 |  |
| 2 | 294 | 298 | 295 | 296 | 308 | 232 |  |
| 3 | 320 | 327 | 319 | 323 | 327 | 296 |  |
| 4 | 346 | 347 | 345 | 362 | 339 | 313 |  |
| 5 | 354 | 354 | 354 | 373 | 345 | 340 |  |
| 6 | 360 | 361 | 359 | 372 | 348 | 348 |  |
| 7 | 369 | 371 | 369 | 394 | 361 | 353 |  |
| 8 | 379 | 379 | 378 | 391 | 377 | 361 |  |
| 9 | 384 | 385 | 384 | 393 | 386 | 370 |  |
| 10 | 388 | 390 | 387 | 400 | 390 | 378 |  |
| 11 | 398 | 402 | 396 | 410 | 408 | 382 |  |
| 12 | 411 | 416 | 405 | 393 | 417 | 384 |  |
| 13 | 427 | 440 | 418 |  | 429 | 390 |  |
| 14 |  |  | 372 |  |  | 391 |  |
| $15+$ |  |  |  |  | 391 |  |  |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017 (cont.).

## Quarter 2

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | 3.6 | 3.c | 3.d | 4.a | 4.6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 230 |  |  | 218 |  | 215 | 215 | 250 | 264 |
| 2 | 282 |  |  | 311 |  |  |  | 311 | 313 |
| 3 | 315 |  |  | 307 | 304 | 310 | 304 | 325 | 315 |
| 4 | 313 |  |  | 336 | 336 | 334 | 336 | 339 | 336 |
| 5 | 329 |  |  | 342 | 340 | 340 | 340 | 348 | 338 |
| 6 | 338 |  |  | 351 | 349 | 351 | 349 | 356 | 347 |
| 7 | 348 |  |  | 354 | 351 | 355 | 351 | 363 | 352 |
| 8 | 355 |  |  | 369 | 369 | 366 | 368 | 371 | 367 |
| 9 | 364 |  |  | 369 | 367 | 367 | 367 | 376 | 368 |
| 10 | 374 |  |  | 385 | 388 | 375 | 386 | 382 | 385 |
| 11 | 383 |  |  | 379 |  | 373 | 373 | 386 | 394 |
| 12 | 380 |  |  | 383 | 380 | 383 | 380 | 390 | 382 |
| 13 | 423 |  |  | 382 | 380 | 380 | 380 | 394 | 381 |
| 14 | 436 |  |  | 365 | 360 | 420 | 362 | 389 | 361 |
| 15+ | 408 |  |  | 408 |  |  |  | 408 | 412 |
| AGE | 4.c | $5 . \mathrm{a}$ | 5.b | $6 . \mathrm{a}$ | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 269 |  |  | 268 |  | 235 |  |  | 269 |
| 2 | 311 | 334 | 334 | 315 | 349 | 271 | 305 | 349 | 311 |
| 3 | 316 | 341 | 332 | 309 | 369 | 290 | 317 | 369 | 318 |
| 4 | 333 | 345 | 348 | 339 | 380 | 304 | 318 | 380 | 333 |
| 5 | 343 | 352 | 353 | 350 | 387 | 330 | 344 | 387 | 340 |
| 6 | 348 | 356 | 359 | 352 | 393 | 345 | 344 | 392 | 346 |
| 7 | 357 | 360 | 363 | 359 | 405 | 349 | 364 | 403 | 360 |
| 8 | 366 | 373 | 371 | 373 | 423 | 349 | 369 | 417 | 360 |
| 9 | 379 | 372 | 376 | 384 | 433 | 380 | 374 | 417 | 372 |
| 10 | 388 | 381 | 380 | 392 | 437 | 382 | 381 | 417 | 375 |
| 11 | 395 | 385 | 384 | 394 | 445 | 395 | 371 | 419 | 396 |
| 12 | 388 | 378 | 396 | 391 | 445 |  | 401 | 445 | 404 |
| 13 | 384 |  | 420 | 395 |  |  | 406 |  | 409 |
| 14 | 367 |  |  | 377 |  |  | 420 |  | 407 |
| 15+ | 413 |  |  | 397 |  |  |  |  | 413 |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017 (cont.).

Quarter 2


Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017 (cont.).

## Quarter 3

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | $3 . \mathrm{b}$ | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 248 |  |  | 288 | 280 | 282 | 294 | 293 | 291 |
| 2 | 329 |  |  | 322 | 311 | 321 | 319 | 311 | 317 |
| 3 | 323 | 350 | 350 | 334 | 328 | 334 | 334 | 325 | 320 |
| 4 | 328 | 355 | 355 | 350 | 350 | 352 | 353 | 341 | 350 |
| 5 | 345 | 362 | 362 | 354 | 348 | 357 | 359 | 348 | 361 |
| 6 | 353 | 370 | 370 | 360 | 346 | 360 | 369 | 356 | 372 |
| 7 | 356 | 387 | 387 | 367 | 353 | 368 | 378 | 362 | 376 |
| 8 | 367 | 397 | 397 | 376 | 365 | 378 | 402 | 368 | 381 |
| 9 | 374 | 392 | 392 | 379 | 370 | 381 | 382 | 373 | 378 |
| 10 | 379 | 405 | 405 | 380 | 380 | 385 | 380 | 378 | 382 |
| 11 | 384 |  |  | 386 | 390 | 390 | 390 | 383 | 391 |
| 12 | 389 |  |  | 392 | 398 | 394 |  | 387 | 389 |
| 13 | 393 |  |  | 402 | 408 | 415 |  | 391 | 387 |
| 14 | 399 |  |  | 389 |  | 385 |  | 396 | 370 |
| 15+ | 408 |  |  | 410 |  |  |  | 410 | 413 |
| AGE | 4.c | 5.a | 5.b | $6 . a$ | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 272 |  |  | 284 |  | 235 | 284 |  | 270 |
| 2 | 311 | 338 | 338 | 316 | 335 | 298 | 325 | 335 | 311 |
| 3 | 317 | 338 | 338 | 335 | 335 | 295 | 335 | 335 | 318 |
| 4 | 334 | 348 | 348 | 351 | 335 | 308 | 350 | 335 | 333 |
| 5 | 339 | 354 | 354 | 351 | 354 | 346 | 354 |  | 340 |
| 6 | 346 | 358 | 358 | 352 | 358 | 349 | 361 |  | 346 |
| 7 | 362 | 364 | 364 | 359 | 364 | 357 | 393 | 395 | 361 |
| 8 | 360 | 373 | 373 | 375 | 373 | 371 | 394 | 395 | 360 |
| 9 | 370 | 376 | 376 | 385 | 376 | 384 | 394 | 395 | 370 |
| 10 | 371 | 382 | 382 | 395 | 382 | 395 | 413 | 415 | 371 |
| 11 | 396 | 387 | 387 | 395 | 387 | 395 | 406 | 406 | 397 |
| 12 | 404 | 393 | 393 | 396 | 393 |  | 408 | 408 | 405 |
| 13 | 414 | 420 | 420 | 411 | 420 |  | 410 | 410 | 415 |
| 14 | 415 |  |  | 385 |  |  | 420 | 420 | 415 |
| 15+ | 413 |  |  |  |  |  |  |  | 413 |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017 (cont.).

Quarter 3

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  | 141 | 141 |  |
| 1 | 280 | 235 | 284 |  |  | 276 | 273 |  |
| 2 | 285 | 266 | 335 | 349 | 335 | 335 | 324 | 317 |
| 3 | 306 | 295 | 335 | 369 | 335 | 335 | 340 | 339 |
| 4 | 317 | 308 | 336 | 380 | 335 | 335 | 350 | 351 |
| 5 | 327 | 320 | 354 | 387 |  | 371 | 370 |  |
| 6 | 339 | 327 | 361 | 393 |  | 382 | 379 |  |
| 7 | 376 | 334 | 380 | 395 | 395 |  | 378 | 379 |
| 8 | 382 | 350 | 388 | 395 | 395 |  | 385 | 385 |
| 9 | 383 | 345 | 388 | 395 | 395 |  |  |  |
| 10 | 402 | 385 | 400 | 415 | 415 |  | 405 | 405 |
| 11 | 406 | 406 | 406 | 406 | 406 |  |  |  |
| 12 | 408 | 408 | 408 | 408 | 408 |  | 429 | 428 |
| 13 | 410 | 410 | 410 | 410 | 410 |  |  |  |
| 14 | 420 | 420 | 420 | 420 | 420 |  |  |  |
| $15+$ | 413 |  |  |  |  |  |  |  |


| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b | All |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 196 |  |  | 192 | 211 |  |  | 145 |
| 1 | 260 | 271 | 289 | 244 | 252 | 352 | 352 | 312 |
| 2 | 313 | 335 | 305 | 324 | 309 | 336 | 336 | 328 |
| 3 | 331 | 340 | 323 | 347 | 333 | 357 | 357 | 330 |
| 4 | 343 | 352 | 341 | 367 | 349 | 372 | 371 | 337 |
| 5 | 368 | 366 | 352 | 397 | 362 | 383 | 381 | 350 |
| 6 | 381 | 381 | 359 |  | 371 | 400 | 380 | 355 |
| 7 | 381 | 395 | 359 |  | 376 | 422 | 367 | 358 |
| 8 | 387 | 394 | 376 |  | 385 |  | 374 | 369 |
| 9 |  |  | 375 |  |  |  | 376 | 375 |
| 10 | 400 | 395 |  |  |  |  | 382 | 380 |
| 11 |  |  |  |  |  |  | 387 | 385 |
| 12 | 415 | 415 |  |  |  | 394 | 390 |  |
| 13 |  |  |  |  |  |  | 393 |  |
| 14 |  |  |  |  |  |  | 399 |  |
| $15+$ |  |  |  |  |  |  | 308 |  |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017 (cont.).

## Quarter 4

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | $3 . \mathrm{b}$ | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 299 | 335 | 335 | 285 | 283 | 282 | 288 | 288 | 289 |
| 2 | 328 | 335 | 335 | 324 | 312 | 321 | 325 | 317 | 312 |
| 3 | 328 | 330 | 330 | 335 | 329 | 334 | 336 | 329 | 323 |
| 4 | 339 | 355 | 355 | 351 | 353 | 352 | 351 | 348 | 346 |
| 5 | 350 | 367 | 367 | 355 | 354 | 357 | 355 | 354 | 354 |
| 6 | 356 | 372 | 372 | 361 | 353 | 360 | 364 | 360 | 364 |
| 7 | 362 | 373 | 373 | 368 | 358 | 368 | 370 | 366 | 367 |
| 8 | 368 | 365 | 365 | 379 | 371 | 378 | 383 | 374 | 372 |
| 9 | 373 | 385 | 385 | 382 | 372 | 381 | 382 | 377 | 375 |
| 10 | 379 | 395 | 395 | 382 | 380 | 385 | 379 | 379 | 375 |
| 11 | 383 | 395 | 395 | 388 | 390 | 390 | 387 | 386 | 391 |
| 12 | 388 | 395 | 395 | 395 |  | 394 | 398 | 392 | 396 |
| 13 | 392 |  |  | 410 |  | 415 | 408 | 400 | 405 |
| 14 | 395 |  |  | 385 |  | 385 |  | 392 | 408 |
| 15+ | 410 |  |  | 410 |  |  |  | 411 | 413 |
| AGE | 4.c | 5.a | 5.b | $6 . a$ | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 269 |  | 280 | 286 |  | 247 | 284 |  | 274 |
| 2 | 312 | 341 | 310 | 313 |  | 269 | 324 |  | 310 |
| 3 | 316 | 332 | 328 | 328 |  | 300 | 335 |  | 318 |
| 4 | 332 | 347 | 340 | 355 |  | 320 | 351 |  | 333 |
| 5 | 337 | 354 | 347 | 354 |  | 343 | 354 |  | 340 |
| 6 | 344 | 359 | 346 | 349 |  | 355 | 361 |  | 347 |
| 7 | 361 | 362 | 352 | 366 |  | 362 | 394 |  | 360 |
| 8 | 357 | 370 | 364 | 377 |  | 373 | 395 |  | 360 |
| 9 | 369 | 375 | 370 | 379 |  | 373 | 395 |  | 371 |
| 10 | 371 | 379 | 380 | 391 |  | 383 | 415 |  | 372 |
| 11 | 397 | 383 | 390 | 405 |  | 390 | 406 |  | 397 |
| 12 | 405 | 395 |  | 408 |  | 394 | 408 |  | 405 |
| 13 | 415 | 420 |  | 410 |  | 415 | 410 |  | 415 |
| 14 | 415 |  |  | 420 |  | 385 | 420 |  | 415 |
| 15+ | 413 |  |  |  |  |  |  |  | 413 |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (mm) -at-age by area for 2017 (cont.).
Quarter 4

| Age | 7.e | 7.f | $7 . \mathrm{g}$ | 7.h | 7.j | 7.k | 8.a | 8.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  | 141 | 141 |
| 1 | 274 | 220 | 283 |  |  |  | 274 | 274 |
| 2 | 282 | 264 | 325 | 335 | 335 |  | 294 | 301 |
| 3 | 308 | 284 | 334 | 335 | 335 |  | 329 | 335 |
| 4 | 321 | 307 | 351 | 335 | 335 |  | 350 | 352 |
| 5 | 339 | 320 | 357 |  |  |  | 368 | 369 |
| 6 | 351 | 330 | 361 |  |  |  | 373 | 376 |
| 7 | 371 | 338 | 371 | 395 | 395 |  | 379 | 379 |
| 8 | 371 | 355 | 380 | 395 | 395 |  | 386 | 386 |
| 9 | 390 | 347 | 383 | 395 | 395 |  |  |  |
| 10 | 390 | 381 | 388 | 415 | 415 |  | 404 | 404 |
| 11 | 406 | 406 | 401 | 406 | 406 |  |  |  |
| 12 | 408 | 408 | 406 | 408 | 408 |  | 419 | 419 |
| 13 | 410 | 410 | 410 | 410 | 410 |  |  |  |
| 14 | 420 | 420 | 414 | 420 | 420 |  |  |  |
| 15+ | 413.0 |  |  |  |  |  |  |  |
| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b | All |
| 0 | 181 |  |  | 184 | 225 |  |  | 143 |
| 1 | 264 | 286 |  | 248 | 258 |  |  | 288 |
| 2 | 295 | 324 |  | 309 | 283 |  |  | 316 |
| 3 | 317 | 335 |  | 348 | 325 |  |  | 329 |
| 4 | 327 | 348 |  | 376 | 345 |  |  | 347 |
| 5 | 368 | 367 |  | 368 | 363 |  |  | 353 |
| 6 | 375 | 376 |  | 371 | 368 |  |  | 359 |
| 7 | 377 | 381 |  | 362 | 376 |  |  | 365 |
| 8 | 385 | 385 |  | 385 | 385 |  |  | 372 |
| 9 |  |  |  |  |  |  |  | 376 |
| 10 | 404 | 405 |  |  |  |  |  | 380 |
| 11 |  |  |  |  |  |  |  | 385 |
| 12 |  | 415 |  |  |  |  |  | 390 |
| 13 |  |  |  |  |  |  |  | 397 |
| 14 |  |  |  |  |  |  |  | 396 |
| 15+ |  |  |  |  |  |  |  | 410 |

Table 8.5.1.2. NE Atlantic Mackerel. Percentage length composition in catches by country and fleet in 2017. Zeros represent values $<1 \%$. Handline Fleet. UKE=UK England and Wales.

|  | UKE LINES |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $7 . \mathrm{E}$ |  |  |  | 7.5 |  |  |  |
| LENGTH См | Q1 | Q2 | Q3 | Q4 | Q1 | Q2 | Q3 | Q4 |
| 15 |  |  |  |  |  |  |  |  |
| 16 |  |  |  |  |  |  |  |  |
| 17 |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  | 0\% |
| 19 |  |  |  |  |  |  |  | 0\% |
| 20 |  |  |  |  |  |  |  | 0\% |
| 21 | 0\% |  | 0\% |  |  |  |  | 0\% |
| 22 | 0\% |  |  | 0\% | 0\% |  |  | 1\% |
| 23 | 0\% |  | 0\% | 1\% | 0\% | 0\% | 0\% | 1\% |
| 24 | 0\% |  | 0\% | 2\% | 0\% | 1\% | 2\% | 1\% |
| 25 | 0\% |  | 3\% | 6\% | 1\% | 2\% | 4\% | 8\% |
| 26 | 3\% | 1\% | 13\% | 11\% | 2\% | 5\% | 3\% | 14\% |
| 27 | 5\% | 5\% | 13\% | 17\% | 4\% | 8\% | 7\% | 13\% |
| 28 | 8\% | 20\% | 7\% | 14\% | 11\% | 8\% | 6\% | 6\% |
| 29 | 20\% | 33\% | 9\% | 4\% | 30\% | 15\% | 14\% | 7\% |
| 30 | 35\% | 22\% | 12\% | 2\% | 29\% | 13\% | 20\% | 10\% |
| 31 | 16\% | 10\% | 14\% | 4\% | 13\% | 8\% | 21\% | 15\% |
| 32 | 8\% | 4\% | 13\% | 6\% | 5\% | 6\% | 10\% | 11\% |
| 33 | 3\% | 1\% | 6\% | 6\% | 2\% | 6\% | 8\% | 6\% |
| 34 | 1\% | 1\% | 3\% | 5\% | 1\% | 9\% | 3\% | 2\% |
| 35 | 0\% | 1\% | 3\% | 3\% | 0\% | 6\% | 1\% | 1\% |
| 36 | 0\% | 1\% | 1\% | 5\% | 0\% | 4\% | 0\% | 1\% |
| 37 | 0\% | 0\% | 1\% | 5\% | 0\% | 4\% | 0\% | 0\% |
| 38 | 0\% | 0\% | 1\% | 2\% | 0\% | 2\% | 0\% | 0\% |
| 39 | 0\% | 0\% | 0\% | 3\% |  | 1\% |  | 0\% |
| 40 | 0\% | 0\% | 0\% | 2\% | 0\% | 1\% | 0\% |  |
| 41 | 0\% |  | 0\% | 0\% |  |  |  |  |
| 42 | 0\% | 0\% |  | 1\% |  | 0\% |  |  |
| 43 | 0\% | 0\% | 0\% | 1\% |  |  |  |  |
| 44 |  |  |  |  |  |  |  |  |
| 45 |  |  |  |  |  |  |  |  |
| 46 |  |  |  |  |  |  |  |  |
| 47 |  |  |  |  |  |  |  |  |

Table 8.5.1.2. NE Atlantic Mackerel. Percentage length composition in catches by country and fleet in 2017. Zeros represent values $<\mathbf{1 \%}$ (cont.). Southern Fleets. ES=Spain.

|  | ES Purse Seine |  |  |  | ES Trawl |  |  |  | ES Artisanal |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LENGTH <br> CM | Q1 | Q2 | Q3 | Q4 | Q1 | Q2 | Q3 | Q4 | Q1 | Q2 | Q3 | Q4 |
| 12 |  |  |  |  |  |  |  |  |  |  |  |  |
| 13 |  |  |  |  |  |  |  |  |  |  |  |  |
| 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 15 |  |  |  |  |  |  |  |  |  |  |  |  |
| 16 |  |  |  |  |  |  |  |  |  |  |  |  |
| 17 |  |  |  |  |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  | 0\% |  |  |  |  |  |  |  |
| 19 |  |  |  |  | 0\% | 0\% | 4\% |  |  |  |  |  |
| 20 |  |  |  |  | 0\% | 0\% | 13\% |  |  |  |  |  |
| 21 |  |  |  |  | 0\% | 0\% | 20\% |  |  |  |  |  |
| 22 |  |  |  | 3\% | 0\% | 0\% | 8\% | 2\% |  |  |  |  |
| 23 |  |  |  | 2\% |  | 0\% | 2\% | 3\% |  |  |  |  |
| 24 |  |  | 0\% | 10\% |  | 0\% | 1\% | 11\% |  |  |  |  |
| 25 |  | 0\% | 4\% | 23\% |  | 1\% | 1\% | 18\% |  |  |  |  |
| 26 |  | 1\% | 9\% | 20\% | 2\% | 0\% | 2\% | 10\% |  |  | 1\% |  |
| 27 |  | 2\% | 7\% | 9\% | 6\% | 1\% | 1\% | 14\% |  |  | 2\% |  |
| 28 | 0\% | 6\% | 3\% | 2\% | 9\% | 2\% | 1\% | 6\% |  |  | 2\% |  |
| 29 | 0\% | 11\% | 1\% | 3\% | 6\% | 3\% | 2\% | 8\% | 0\% |  | 5\% | 2\% |
| 30 | 0\% | 9\% | 3\% | 6\% | 11\% | 7\% | 7\% | 7\% | 0\% |  | 5\% | 5\% |
| 31 | 1\% | 3\% | 7\% | 5\% | 11\% | 10\% | 10\% | 5\% | 0\% | 0\% | 19\% | 16\% |
| 32 | 2\% | 3\% | 6\% | 8\% | 5\% | 11\% | 7\% | 6\% | 1\% | 0\% | 32\% | 35\% |
| 33 | 7\% | 5\% | 8\% | 4\% | 3\% | 11\% | 4\% | 2\% | 5\% | 3\% | 14\% | 20\% |
| 34 | 18\% | 9\% | 14\% | 2\% | 10\% | 10\% | 4\% | 2\% | 17\% | 12\% | 10\% | 9\% |
| 35 | 23\% | 12\% | 24\% | 2\% | 12\% | 10\% | 5\% | 2\% | 21\% | 16\% | 6\% | 7\% |
| 36 | 18\% | 12\% | 9\% | 3\% | 10\% | 9\% | 5\% | 3\% | 18\% | 18\% | 3\% | 7\% |
| 37 | 14\% | 11\% | 4\% |  | 7\% | 10\% | 3\% | 2\% | 17\% | 20\% |  |  |
| 38 | 10\% | 7\% | 0\% |  | 3\% | 9\% | 1\% | 0\% | 12\% | 16\% |  |  |
| 39 | 4\% | 4\% | 0\% |  | 2\% | 2\% | 0\% |  | 6\% | 9\% |  |  |
| 40 | 2\% | 3\% |  |  | 0\% | 1\% |  |  | 2\% | 4\% |  |  |
| 41 | 0\% | 1\% |  |  | 0\% | 1\% |  |  | 1\% | 1\% |  |  |
| 42 | 0\% | 0\% |  |  | 0\% | 1\% |  |  | 0\% | 1\% |  |  |
| 43 | 0\% | 0\% |  |  | 0\% | 0\% |  |  | 0\% | 0\% |  |  |
| 44 |  | 0\% |  |  |  |  |  |  |  | 0\% |  |  |
| 45 |  |  |  |  |  | 0\% |  |  |  |  |  |  |
| 46 |  |  |  |  |  |  |  |  |  |  |  |  |
| 47 |  |  |  |  |  |  |  |  |  |  |  |  |
| 49 |  |  |  |  |  |  |  |  |  |  |  |  |

Table 8.5.1.2. NE Atlantic Mackerel. Percentage length composition in catches by country and fleet in 2017. Zeros represent values $<1 \%$ (cont.). Southern Fleets (cont.). BQ=Basque, PT=Portugal.

|  | BQ Purse Seine |  |  |  | BQ Artisanal |  |  |  | $\begin{gathered} \hline \text { BQ } \\ \text { Trawl } \end{gathered}$ |  |  | PT All |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \text { LENG } \\ \text { TH } \\ \text { CM } \end{gathered}$ | Q1 | Q2 | Q 3 | Q4 | Q1 | Q2 | Q3 | Q4 | Q1 | Q2 | Q4 | Q1 | Q2 | Q3 | Q4 |
| 11 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 12 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 13 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 14 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 15 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 16 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 17 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 19 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 20 |  |  |  |  |  |  |  |  | 0\% |  |  |  |  |  |  |
| 21 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 22 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 23 |  |  |  |  |  |  |  |  | 0\% |  |  |  |  |  | 3\% |
| 24 |  |  |  |  | 1\% |  |  |  | 1\% |  |  |  |  |  | $\begin{aligned} & 13 \\ & \% \end{aligned}$ |
| 25 |  | 0\% |  |  | 2\% |  |  |  | 3\% | 0\% |  |  |  | 0\% | $\begin{aligned} & 11 \\ & \% \end{aligned}$ |
| 26 |  |  |  | 3\% | 1\% |  |  |  | 3\% | 0\% |  |  | 0\% | 0\% | 9\% |
| 27 |  |  |  | 1\% | 1\% |  |  |  | 7\% | 0\% |  |  | 0\% | 0\% | 1\% |
| 28 |  |  |  | 2\% | 0\% |  |  |  | $\begin{aligned} & 10 \\ & \% \\ & \hline \end{aligned}$ | 1\% |  | 5\% | 1\% | 0\% | 2\% |
| 29 | 0\% |  |  | 1\% | 0\% |  |  |  | $\begin{aligned} & 22 \\ & \% \end{aligned}$ | 4\% | $\begin{aligned} & 13 \\ & \% \end{aligned}$ | 4\% | 1\% | 1\% | 2\% |
| 30 | 0\% | 1\% |  | 5\% | 0\% |  |  |  | $\begin{aligned} & 22 \\ & \% \end{aligned}$ | 5\% | 7\% | 6\% | 5\% | 1\% | 2\% |
| 31 | 0\% | 1\% |  | 8\% | 0\% | 0\% |  |  | $\begin{aligned} & 21 \\ & \% \\ & \hline \end{aligned}$ | 5\% | $\begin{aligned} & 20 \\ & \% \\ & \hline \end{aligned}$ | $\begin{aligned} & 11 \\ & \% \\ & \hline \end{aligned}$ | 8\% | 6\% | 2\% |
| 32 | 2\% | 4\% |  | $\begin{aligned} & 15 \\ & \% \\ & \hline \end{aligned}$ | 2\% | 2\% | $\begin{aligned} & 19 \\ & \% \\ & \hline \end{aligned}$ |  | 7\% | 2\% | $\begin{aligned} & 13 \\ & \% \\ & \hline \end{aligned}$ | 7\% | $\begin{aligned} & 15 \\ & \% \\ & \hline \end{aligned}$ | $\begin{aligned} & 17 \\ & \% \\ & \hline \end{aligned}$ | 5\% |
| 33 | 7\% | 9\% |  | $\begin{aligned} & 27 \\ & \% \end{aligned}$ | 5\% | 4\% | $\begin{aligned} & 18 \\ & \% \end{aligned}$ | $\begin{aligned} & 50 \\ & \% \end{aligned}$ | 2\% | 1\% | $\begin{aligned} & 27 \\ & \% \end{aligned}$ | 6\% | $\begin{aligned} & 12 \\ & \% \end{aligned}$ | $\begin{aligned} & 17 \\ & \% \end{aligned}$ | 8\% |
| 34 | $\begin{array}{\|l\|} \hline 15 \\ \% \end{array}$ | $\begin{aligned} & 17 \\ & \% \end{aligned}$ |  | $\begin{aligned} & 17 \\ & \% \end{aligned}$ | $\begin{aligned} & 13 \\ & \% \end{aligned}$ | $\begin{aligned} & 13 \\ & \% \end{aligned}$ | $\begin{aligned} & 26 \\ & \% \end{aligned}$ | $\begin{aligned} & 50 \\ & \% \end{aligned}$ | 1\% | 1\% | 7\% | 5\% | 7\% | $\begin{aligned} & 17 \\ & \% \end{aligned}$ | 8\% |
| 35 | $\begin{aligned} & \hline 19 \\ & \% \end{aligned}$ | $\begin{aligned} & 21 \\ & \% \end{aligned}$ |  | $\begin{aligned} & 11 \\ & \% \end{aligned}$ | $\begin{aligned} & 17 \\ & \% \end{aligned}$ | $\begin{aligned} & 22 \\ & \% \end{aligned}$ | $\begin{aligned} & 24 \\ & \% \end{aligned}$ |  |  | 2\% |  | 4\% | $\begin{aligned} & 15 \\ & \% \end{aligned}$ | $\begin{aligned} & 17 \\ & \% \end{aligned}$ | 9\% |
| 36 | $\begin{array}{\|l\|} \hline 21 \\ \% \end{array}$ | $\begin{aligned} & 15 \\ & \% \end{aligned}$ |  | 5\% | $\begin{aligned} & 18 \\ & \% \end{aligned}$ | $\begin{aligned} & 21 \\ & \% \end{aligned}$ | $\begin{aligned} & 13 \\ & \% \end{aligned}$ |  |  | 2\% | 7\% | 7\% | $\begin{aligned} & 10 \\ & \% \end{aligned}$ | $\begin{aligned} & 10 \\ & \% \end{aligned}$ | 6\% |
| 37 | $\begin{array}{\|l\|} \hline 15 \\ \% \\ \hline \end{array}$ | $\begin{aligned} & 13 \\ & \% \end{aligned}$ |  | 2\% | $\begin{aligned} & 17 \\ & \% \end{aligned}$ | $\begin{aligned} & 17 \\ & \% \end{aligned}$ |  |  |  | 3\% |  | 9\% | 5\% | 4\% | 8\% |
| 38 | $\begin{aligned} & 11 \\ & \% \end{aligned}$ | $\begin{aligned} & 11 \\ & \% \end{aligned}$ |  | 3\% | $\begin{aligned} & 12 \\ & \% \end{aligned}$ | $\begin{aligned} & 12 \\ & \% \end{aligned}$ |  |  |  | $\begin{aligned} & 13 \\ & \% \end{aligned}$ |  | $\begin{aligned} & 13 \\ & \% \end{aligned}$ | 7\% | 4\% | 8\% |
| 39 | 5\% | 5\% |  |  | 6\% | 5\% |  |  |  | 8\% |  | $\begin{aligned} & 12 \\ & \% \\ & \hline \end{aligned}$ | 6\% | 1\% | 3\% |
| 40 | 2\% | 2\% |  |  | 2\% | 3\% |  |  |  | $\begin{aligned} & 18 \\ & \% \end{aligned}$ | 7\% | 6\% | 4\% | 1\% | 1\% |


|  | BQ Purse Seine |  |  |  | BQ Artisanal |  |  |  | $\begin{gathered} \text { BQ } \\ \text { Trawl } \end{gathered}$ |  |  | PT All |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \text { LENG } \\ \text { TH } \\ \text { CM } \\ \hline \end{gathered}$ | Q1 | Q2 | $\begin{gathered} \mathrm{Q} \\ 3 \end{gathered}$ | Q4 | Q1 | Q2 | Q3 | Q4 | Q1 | Q2 | Q4 | Q1 | Q2 | Q3 | Q4 |
| 41 | 1\% | 1\% |  |  | 1\% | 1\% |  |  |  | $\begin{aligned} & 18 \\ & \% \\ & \hline \end{aligned}$ |  | 3\% | 2\% | 0\% | 0\% |
| 42 | 0\% | 0\% |  |  | 0\% | 0\% |  |  |  | 0\% |  | 1\% | 0\% | 2\% | 0\% |
| 43 | 0\% |  |  |  | 0\% | 0\% |  |  |  | $\begin{aligned} & 18 \\ & \% \end{aligned}$ |  | 1\% | 0\% |  |  |
| 44 | 0\% |  |  |  | 0\% |  |  |  |  |  |  | 0\% |  |  |  |
| 45 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0\% |
| 46 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 47 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 49 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 8.5.1.2. NE Atlantic Mackerel. Percentage length composition in catches by country and fleet in 2017. Zeros represent values $<1 \%$ (cont.). Pelagic Trawl Fleets. IE=Ireland, UKS=UK Scotland, IS=Iceland, DK=Denmark.

|  | IE |  |  | UKS |  | IS |  |  | DK <br> 4.A |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 4.A | $6 . \mathrm{A}$ | 7.8 | 4.A | 6.A | 2.A, 5.A, 14.B |  |  |  |
| Length <br> см | Q4 | Q1 | Q1 | Q4 | Q1 | Q2 | Q3 | Q4 | Q4 |
| 15 |  |  |  |  |  |  |  |  |  |
| 16 |  | 0\% |  |  |  |  |  |  |  |
| 17 |  |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |  |  |
| 19 |  |  |  |  |  |  |  |  |  |
| 20 |  | 0\% |  |  |  |  |  |  |  |
| 21 |  |  |  |  |  |  |  |  |  |
| 22 |  |  |  |  | 0\% |  |  |  |  |
| 23 |  | 0\% |  |  |  |  |  |  |  |
| 24 | 0\% | 0\% |  |  | 0\% |  |  |  |  |
| 25 | 0\% |  |  |  |  |  |  |  |  |
| 26 |  |  |  | 0\% |  |  | 0\% |  |  |
| 27 | 1\% | 0\% |  | 0\% | 0\% |  | 0 |  |  |
| 28 | 1\% | 0\% |  | 1\% | 0\% |  | 0\% |  |  |
| 29 | 1\% | 0\% | 0\% | 1\% | 1\% |  | 0\% | 0\% | 0\% |
| 30 | 0\% | 1\% | 1\% | 1\% | 2\% | 0\% | 0\% | 1\% | 3\% |
| 31 | 2\% | 2\% | 1\% | 3\% | 2\% | 1\% | 3\% | 6\% | 3\% |
| 32 | 6\% | 1\% | 2\% | 8\% | 3\% | 5\% | 5\% | 8\% | 3\% |
| 33 | 10\% | 6\% | 6\% | 10\% | 12\% | 17\% | 13\% | 15\% | 5\% |
| 34 | 12\% | 18\% | 16\% | 18\% | 20\% | 27\% | 24\% | 24\% | 15\% |
| 35 | 21\% | 19\% | 20\% | 21\% | 18\% | 22\% | 21\% | 20\% | 20\% |
| 36 | 19\% | 15\% | 16\% | 17\% | 14\% | 13\% | 16\% | 14\% | 15\% |
| 37 | 13\% | 15\% | 18\% | 11\% | 15\% | 9\% | 11\% | 7\% | 11\% |
| 38 | 9\% | 12\% | 12\% | 7\% | 8\% | 4\% | 5\% | 3\% | 12\% |
| 39 | 4\% | 7\% | 5\% | 3\% | 3\% | 1\% | 2\% | 2\% | 7\% |
| 40 | 1\% | 3\% | 2\% | 1\% | 1\% | 0\% | 1\% | 0\% | 3\% |
| 41 | 1\% | 1\% | 1\% | 0\% | 0\% |  | 0\% | 0\% | 1\% |
| 42 | 0\% | 0\% | 0\% | 0\% | 0\% |  |  |  | 0\% |
| 43 | 0\% | 0\% | 0\% |  | 0\% |  | 0\% |  | 0\% |
| 44 |  |  |  | 0\% | 0\% |  |  |  | 0\% |
| 45 |  |  |  |  |  |  |  |  |  |

Table 8.5.1.2. NE Atlantic Mackerel. Percentage length composition in catches by country and fleet in 2017. Zeros represent values $<\mathbf{1 \%}$ (cont.). Freezer Trawlers. NL=The Netherlands, DE=Germany, RU= Russia.

|  | NL |  |  |  |  | DE |  |  | RU |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2.А,4.А,4.В,6.А,7.В,7.С |  |  |  | $6 . A$ | $4 . \mathrm{A}$ | 2.A | 7.в | 2.A | 2.A |
| $\begin{gathered} \text { LENGTH } \\ \text { CM } \end{gathered}$ | Q1 | Q2 | Q3 | Q4 | Q1 | Q3 | Q4 | Q1 | Q3 | Q4 |
| 15 |  |  |  |  |  |  |  |  |  |  |
| 16 | 0\% |  |  |  |  |  |  |  |  |  |
| 17 | 1\% |  |  |  |  |  |  |  |  |  |
| 18 | 1\% |  |  |  |  |  |  |  |  |  |
| 19 | 0\% |  |  |  |  |  |  |  |  |  |
| 20 | 0\% |  |  |  |  |  |  |  |  |  |
| 21 |  |  |  |  | 0\% | 0\% | 0\% |  | 0\% |  |
| 22 |  |  |  |  | 0\% | 0\% | 0\% |  | 0\% |  |
| 23 |  |  |  |  | 0\% | 0\% | 0\% |  | 0\% | 0\% |
| 24 |  |  |  |  | 0\% | 0\% | 0\% |  |  | 0\% |
| 25 |  |  |  |  | 0\% | 0\% | 0\% |  | 0\% | 0\% |
| 26 | 1\% |  | 0\% |  | 0\% | 0\% | 0\% |  |  | 0\% |
| 27 | 0\% |  |  |  | 1\% | 0\% | 0\% | 0\% | 0\% | 0\% |
| 28 | 1\% |  | 0\% |  | 2\% | 1\% | 1\% | 1\% | 0\% | 0\% |
| 29 | 3\% |  | 2\% |  | 3\% | 5\% | 5\% | 3\% | 0\% | 0\% |
| 30 | 8\% |  | 2\% |  | 3\% | 24\% | 24\% | 3\% | 5\% | 1\% |
| 31 | 6\% | 0.04 | 6\% | 2\% | 2\% | 28\% | 28\% | 4\% | 22\% | 10\% |
| 32 | 2\% | 0\% | 6\% | 9\% | 5\% | 12\% | 12\% | 4\% | 19\% | 21\% |
| 33 | 6\% | 4\% | 4\% | 9\% | 14\% | 10\% | 10\% | 13\% | 10\% | 15\% |
| 34 | 14\% | 12\% | 7\% | 8\% | 17\% | 8\% | 8\% | 22\% | 16\% | 17\% |
| 35 | 18\% | 28\% | 15\% | 16\% | 15\% | 5\% | 5\% | 16\% | 15\% | 18\% |
| 36 | 13\% | 8\% | 14\% | 18\% | 14\% | 3\% | 3\% | 16\% | 7\% | 10\% |
| 37 | 10\% | 8\% | 9\% | 19\% | 12\% | 1\% | 1\% | 12\% | 3\% | 4\% |
| 38 | 10\% | 24\% | 15\% | 8\% | 6\% | 0\% | 0\% | 4\% | 1\% | 2\% |
| 39 | 5\% | 4\% | 10\% | 7\% | 2\% | 0\% | 0\% | 2\% | 1\% | 1\% |
| 40 | 1\% | 8\% | 7\% | 4\% | 1\% | 0\% | 0\% | 1\% | 0\% | 0\% |
| 41 | 2\% |  | 2\% |  | 0\% | 0\% | 0\% | 0\% | 0\% | 0\% |
| 42 | 1\% |  | 0\% |  | 0\% | 0\% | 0\% | 0\% | 0\% | 0\% |
| 43 | 0\% |  | 0\% |  | 0\% |  |  |  | 0\% | 0\% |
| 44 |  |  |  |  |  |  |  |  | 0\% |  |
| 45 |  |  |  |  |  |  |  |  | 0\% | 0\% |
| 46 |  |  |  |  |  |  |  |  |  |  |
| 47 |  |  |  |  |  |  |  |  |  |  |
| 48 |  |  |  |  |  |  |  |  |  |  |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017.
Quarters 1-4

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | 3.b | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 220 | 318 | 318 | 188 | 171 | 176 | 199 | 199 | 191 |
| 2 | 333 | 316 | 316 | 282 | 246 | 272 | 278 | 270 | 269 |
| 3 | 327 | 380 | 306 | 289 | 290 | 309 | 300 | 302 | 259 |
| 4 | 348 | 407 | 386 | 351 | 343 | 368 | 375 | 365 | 330 |
| 5 | 392 | 430 | 424 | 368 | 355 | 384 | 391 | 380 | 345 |
| 6 | 418 | 464 | 442 | 389 | 352 | 397 | 415 | 402 | 363 |
| 7 | 434 | 527 | 465 | 404 | 376 | 429 | 424 | 421 | 366 |
| 8 | 466 | 542 | 462 | 446 | 422 | 467 | 486 | 454 | 412 |
| 9 | 489 | 533 | 494 | 457 | 442 | 481 | 445 | 468 | 416 |
| 10 | 510 | 593 | 581 | 484 | 487 | 497 | 481 | 483 | 475 |
| 11 | 528 | 612 | 612 | 494 | 532 | 514 | 503 | 506 | 534 |
| 12 | 549 | 542 | 542 | 493 | 467 | 531 | 459 | 520 | 463 |
| 13 | 572 |  |  | 482 | 459 | 606 | 455 | 526 | 457 |
| 14 | 604 |  |  | 406 | 376 | 497 | 384 | 431 | 383 |
| 15+ | 639 |  |  | 634 |  |  |  | 614 | 611 |
| AGE | 4.c | 5.a | 5.b | $6 . a$ | $6 . \mathrm{b}$ | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  | 27 |  | 27 |  |  |  |
| 1 | 175 |  | 170 | 192 |  | 111 | 187 | 187 | 179 |
| 2 | 262 | 369 | 318 | 193 | 276 | 177 | 212 | 272 | 260 |
| 3 | 276 | 370 | 256 | 232 | 276 | 214 | 237 | 314 | 272 |
| 4 | 320 | 400 | 352 | 310 | 277 | 261 | 250 | 361 | 314 |
| 5 | 331 | 419 | 388 | 332 | 416 | 304 | 314 | 342 | 338 |
| 6 | 345 | 433 | 399 | 350 | 432 | 318 | 317 | 383 | 357 |
| 7 | 353 | 454 | 399 | 371 | 454 | 332 | 363 | 417 | 392 |
| 8 | 396 | 485 | 439 | 407 | 486 | 376 | 391 | 403 | 414 |
| 9 | 401 | 494 | 454 | 440 | 495 | 401 | 406 | 429 | 434 |
| 10 | 451 | 516 | 490 | 455 | 516 | 444 | 424 | 464 | 449 |
| 11 | 432 | 539 | 537 | 464 | 539 | 400 | 402 | 504 | 522 |
| 12 | 514 | 563 | 478 | 491 | 564 | 476 | 476 | 527 | 526 |
| 13 | 485 | 671 | 452 | 504 | 670 | 470 | 550 | 588 | 513 |
| 14 | 434 |  | 376 | 546 |  | 449 | 664 | 708 | 574 |
| 15+ | 602 |  |  | 530 |  | 434 |  |  | 602 |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017 (cont.).

Quarters 1-4

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  |  | 17 | 17 |
| 1 | 185 | 97 | 187 | 187 | 187 |  | 79 | 49 |
| 2 | 208 | 150 | 274 | 275 | 267 | 289 | 201 | 196 |
| 3 | 238 | 190 | 300 | 301 | 237 | 338 | 267 | 250 |
| 4 | 256 | 218 | 306 | 302 | 280 | 369 | 306 | 298 |
| 5 | 308 | 247 | 338 | 331 | 312 | 389 | 329 | 318 |
| 6 | 330 | 270 | 338 | 330 | 323 | 408 | 349 | 334 |
| 7 | 414 | 299 | 363 | 399 | 356 | 432 | 378 | 365 |
| 8 | 431 | 356 | 405 | 402 | 405 | 426 | 407 | 398 |
| 9 | 453 | 356 | 412 | 411 | 410 | 439 | 431 | 421 |
| 10 | 472 | 460 | 429 | 434 | 429 | 473 | 462 | 435 |
| 11 | 553 | 572 | 456 | 476 | 491 | 516 | 478 | 471 |
| 12 | 608 | 611 | 476 | 605 | 501 | 594 | 583 | 539 |
| 13 | 616 | 617 | 566 | 615 | 584 |  | 630 | 623 |
| 14 | 716 | 719 | 665 | 719 | 692 |  |  |  |
| 15 | 602 |  |  |  |  |  |  |  |

$15+\quad 602$

| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a. $\mathbf{n}$ | 14.a | 14.b | All |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 47 |  | 15 | 54 | 74 |  |  | 18 |
| 1 | 120 | 124 | 46 | 81 | 127 | 431 | 431 | 178 |
| 2 | 178 | 189 | 181 | 260 | 182 | 369 | 369 | 266 |
| 3 | 228 | 251 | 226 | 307 | 243 | 422 | 422 | 312 |
| 4 | 283 | 300 | 283 | 397 | 289 | 481 | 479 | 356 |
| 5 | 305 | 322 | 302 | 421 | 298 | 532 | 523 | 377 |
| 6 | 320 | 341 | 310 | 405 | 311 | 577 | 516 | 397 |
| 7 | 350 | 375 | 322 | 473 | 354 | 644 | 477 | 415 |
| 8 | 374 | 397 | 363 | 464 | 381 |  | 501 | 444 |
| 9 | 387 | 422 | 377 | 483 | 397 |  | 510 | 466 |
| 10 | 400 | 437 | 442 | 510 | 412 |  | 532 | 484 |
| 11 | 444 | 501 | 374 | 539 | 443 |  | 554 | 497 |
| 12 | 482 | 535 | 474 | 483 | 473 |  | 581 | 523 |
| 13 | 549 | 621 | 528 |  | 574 |  |  | 537 |
| 14 |  |  | 424 |  |  |  | 535 |  |
| $15+$ |  |  |  |  |  | 606 |  |  |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017 (cont.).

## Quarter 1

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | 3.b | 3.c | 3.d | 4.a | 4.6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 188 |  |  |  |  | 67 |  | 77 | 106 |
| 2 | 358 |  |  |  |  |  |  | 206 | 199 |
| 3 | 345 |  |  | 215 |  | 232 | 215 | 215 | 211 |
| 4 | 382 |  |  | 300 |  | 298 | 300 | 299 | 296 |
| 5 | 403 |  |  | 312 |  | 316 | 312 | 315 | 326 |
| 6 | 422 |  |  | 339 |  | 353 | 339 | 341 | 347 |
| 7 | 443 |  |  | 348 |  | 365 | 348 | 351 | 361 |
| 8 | 475 |  |  | 409 |  | 403 | 409 | 410 | 419 |
| 9 | 491 |  |  | 400 |  | 406 | 400 | 405 | 419 |
| 10 | 517 |  |  | 479 |  | 437 | 479 | 477 | 475 |
| 11 | 538 |  |  |  |  | 430 |  | 452 | 456 |
| 12 | 552 |  |  | 450 |  | 471 | 450 | 452 | 456 |
| 13 | 581 |  |  | 450 |  | 457 | 450 | 451 | 453 |
| 14 | 599 |  |  | 376 |  | 643 | 376 | 378 | 378 |
| 15+ | 641 |  |  |  |  |  |  | 641 |  |
| AGE | 4.c | 5.a | 5.b | $6 . a$ | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  | 27 |  | 27 |  |  |  |
| 1 | 106 |  |  | 193 |  | 96 | 187 | 187 | 187 |
| 2 | 196 | 358 | 289 | 192 |  | 190 | 212 | 267 | 219 |
| 3 | 205 | 376 | 215 | 232 |  | 220 | 236 | 259 | 206 |
| 4 | 290 | 386 | 300 | 310 |  | 291 | 251 | 357 | 290 |
| 5 | 336 | 406 | 312 | 332 |  | 316 | 314 | 321 | 335 |
| 6 | 358 | 420 | 339 | 350 |  | 332 | 317 | 380 | 357 |
| 7 | 381 | 431 | 348 | 372 |  | 363 | 361 | 416 | 380 |
| 8 | 438 | 475 | 409 | 407 |  | 389 | 389 | 402 | 437 |
| 9 | 437 | 469 | 400 | 441 |  | 432 | 404 | 429 | 437 |
| 10 | 471 | 502 | 479 | 455 |  | 440 | 423 | 463 | 469 |
| 11 | 458 | 515 | 594 | 465 |  | 439 | 392 | 503 | 457 |
| 12 | 474 | 492 | 450 | 491 |  | 476 | 452 | 475 | 474 |
| 13 | 460 |  | 450 | 502 |  | 470 | 522 | 528 | 464 |
| 14 | 423 |  | 376 | 544 |  | 448 | 507 | 507 | 412 |
| 15+ |  |  |  | 530 |  | 434 |  |  |  |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017 (cont.).

Quarter 1

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  | 14 | 14 |  |
| 1 | 186 | 128 | 187 | 187 | 187 |  | 36 | 46 |
| 2 | 245 | 158 | 274 | 272 | 266 |  | 193 | 194 |
| 3 | 209 | 187 | 315 | 259 | 233 |  | 244 | 245 |
| 4 | 224 | 202 | 336 | 242 | 280 | 442 | 291 | 293 |
| 5 | 291 | 221 | 333 | 314 | 312 | 337 | 312 | 315 |
| 6 | 310 | 231 | 327 | 313 | 323 | 401 | 328 | 330 |
| 7 | 358 | 252 | 348 | 373 | 349 | 425 | 358 | 359 |
| 8 | 390 | 311 | 399 | 386 | 397 | 408 | 393 | 394 |
| 9 | 404 | 337 | 408 | 403 | 407 | 433 | 415 | 415 |
| 10 | 429 | 569 | 422 | 423 | 422 | 469 | 426 | 426 |
| 11 | 397 | 387 | 418 | 388 | 411 | 515 | 467 | 466 |
| 12 | 500 | 541 | 434 | 541 | 435 |  | 515 | 511 |
| 13 | 534 | 541 | 517 | 541 | 518 |  | 615 | 615 |
| 14 | 424 |  | 507 |  | 507 |  |  |  |
| $15+$ |  |  |  |  |  |  |  |  |

15+

| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  | 21 |  |  | All |  |
| 1 | 134 | 124 | 58 | 52 | 134 | 14 |  |
| 2 | 175 | 186 | 176 | 200 | 201 | 105 |  |
| 3 | 225 | 249 | 224 | 262 | 239 | 189 |  |
| 4 | 283 | 300 | 284 | 375 | 266 | 230 |  |
| 5 | 304 | 321 | 304 | 409 | 280 | 306 |  |
| 6 | 318 | 341 | 319 | 406 | 289 | 328 |  |
| 7 | 344 | 374 | 346 | 487 | 321 | 346 |  |
| 8 | 369 | 397 | 373 | 473 | 365 | 369 |  |
| 9 | 384 | 420 | 388 | 480 | 391 | 405 |  |
| 10 | 397 | 435 | 405 | 510 | 404 | 436 |  |
| 11 | 428 | 481 | 434 | 549 | 462 | 454 |  |
| 12 | 471 | 536 | 474 | 483 | 490 | 462 |  |
| 13 | 528 | 637 | 528 |  | 532 | 487 |  |
| 14 |  |  | 424 |  |  | 493 |  |
| $15+$ |  |  |  |  | 491 |  |  |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017 (cont.).

## Quarter 2

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | $3 . \mathrm{b}$ | 3.c | 3.d | $4 . \mathrm{a}$ | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 125 |  |  | 86 |  | 67 | 67 | 182 | 166 |
| 2 | 226 |  |  | 261 |  |  |  | 262 | 269 |
| 3 | 301 |  |  | 231 | 215 | 232 | 216 | 312 | 255 |
| 4 | 292 |  |  | 304 | 300 | 298 | 300 | 351 | 308 |
| 5 | 335 |  |  | 329 | 312 | 316 | 312 | 389 | 324 |
| 6 | 371 |  |  | 361 | 339 | 353 | 341 | 420 | 346 |
| 7 | 397 |  |  | 372 | 348 | 365 | 350 | 446 | 357 |
| 8 | 426 |  |  | 421 | 409 | 403 | 408 | 471 | 409 |
| 9 | 428 |  |  | 431 | 400 | 406 | 401 | 496 | 413 |
| 10 | 486 |  |  | 483 | 479 | 437 | 475 | 518 | 475 |
| 11 | 505 |  |  | 478 |  | 430 | 430 | 542 | 536 |
| 12 | 537 |  |  | 475 | 450 | 471 | 452 | 555 | 462 |
| 13 | 684 |  |  | 468 | 450 | 457 | 451 | 569 | 456 |
| 14 | 737 |  |  | 400 | 376 | 643 | 386 | 542 | 382 |
| 15+ | 641 |  |  | 641 |  |  |  | 641 | 611 |
|  |  |  |  |  |  |  |  |  |  |
| AGE | 4.c | $5 . \mathrm{a}$ | 5.b | $6 . a$ | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  | 27 |  |  |  |  |  |
| 1 | 173 |  |  | 163 |  | 100 |  |  | 173 |
| 2 | 259 | 358 | 334 | 198 | 289 | 153 | 214 | 289 | 261 |
| 3 | 275 | 376 | 341 | 217 | 342 | 191 | 239 | 342 | 280 |
| 4 | 317 | 386 | 390 | 301 | 374 | 219 | 244 | 374 | 321 |
| 5 | 325 | 406 | 411 | 306 | 393 | 276 | 314 | 393 | 337 |
| 6 | 335 | 420 | 430 | 311 | 412 | 314 | 315 | 412 | 354 |
| 7 | 334 | 431 | 445 | 318 | 452 | 308 | 378 | 450 | 390 |
| 8 | 389 | 475 | 473 | 378 | 515 | 326 | 403 | 498 | 396 |
| 9 | 389 | 469 | 492 | 385 | 549 | 410 | 420 | 502 | 421 |
| 10 | 453 | 502 | 510 | 450 | 565 | 428 | 436 | 519 | 441 |
| 11 | 403 | 515 | 524 | 382 | 594 | 369 | 407 | 538 | 499 |
| 12 | 491 | 492 | 574 | 483 | 594 |  | 524 | 594 | 566 |
| 13 | 467 |  | 690 | 510 |  |  | 546 |  | 585 |
| 14 | 405 |  |  | 440 |  |  | 713 |  | 578 |
| 15+ | 602 |  |  | 519 |  |  |  |  | 602 |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017 (cont.).
Quarter 2

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  | 14 | 14 |  |
| 1 | 123 | 100 |  |  |  | 34 | 34 |  |
| 2 | 267 | 154 | 286 | 288 | 275 | 289 | 243 | 209 |
| 3 | 257 | 192 | 310 | 334 | 276 | 342 | 281 | 264 |
| 4 | 250 | 220 | 317 | 361 | 287 | 374 | 313 | 304 |
| 5 | 328 | 273 | 334 | 370 | 321 | 393 | 331 | 324 |
| 6 | 336 | 318 | 333 | 376 | 324 | 412 | 351 | 345 |
| 7 | 380 | 315 | 356 | 410 | 446 | 452 | 385 | 381 |
| 8 | 390 | 327 | 397 | 403 | 486 | 515 | 411 | 408 |
| 9 | 406 | 423 | 408 | 417 | 463 | 549 | 436 | 433 |
| 10 | 424 | 423 | 423 | 437 | 497 | 565 | 449 | 446 |
| 11 | 392 | 388 | 407 | 412 | 568 | 594 | 480 | 479 |
| 12 | 516 | 543 | 437 | 549 | 603 | 594 | 499 | 501 |
| 13 | 537 | 541 | 518 | 541 | 616 |  | 634 | 634 |
| 14 | 424 |  | 507 |  | 718 |  |  |  |
| $15+$ | 602 |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |


| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b | All |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  | 14 |  |  | 431 | 62 |  |
| 1 | 97 | 122 | 34 | 68 | 97 | 369 | 234 |  |
| 2 | 183 | 207 | 193 | 248 | 176 | 422 | 279 |  |
| 3 | 230 | 262 | 237 | 264 | 214 | 481 | 296 |  |
| 4 | 281 | 304 | 277 | 360 | 261 | 532 | 327 |  |
| 5 | 306 | 323 | 300 | 372 | 302 | 577 | 352 |  |
| 6 | 324 | 345 | 292 | 400 | 327 | 644 | 378 |  |
| 7 | 357 | 384 | 303 | 459 | 366 |  | 405 |  |
| 8 | 379 | 403 | 338 | 456 | 385 | 412 |  |  |
| 9 | 390 | 435 | 373 | 488 | 397 | 456 |  |  |
| 10 | 402 | 451 | 450 | 509 | 413 |  | 460 |  |
| 11 | 453 | 549 | 369 | 527 | 440 | 519 |  |  |
| 12 | 487 | 522 | 549 | 450 | 468 |  | 537 |  |
| 13 | 552 | 615 |  |  | 580 | 463 |  |  |
| 14 |  |  |  |  |  | 619 |  |  |
| $15+$ |  |  |  |  |  |  |  |  |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017 (cont.).

## Quarter 3

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | $3 . \mathrm{b}$ | $3 . \mathrm{c}$ | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 161 |  |  | 196 | 171 | 177 | 202 | 216 | 202 |
| 2 | 344 |  |  | 281 | 246 | 272 | 275 | 262 | 269 |
| 3 | 333 | 403 | 403 | 316 | 293 | 309 | 324 | 296 | 268 |
| 4 | 353 | 411 | 411 | 368 | 372 | 368 | 389 | 344 | 371 |
| 5 | 397 | 431 | 431 | 380 | 359 | 384 | 406 | 367 | 409 |
| 6 | 423 | 465 | 465 | 402 | 353 | 397 | 440 | 393 | 447 |
| 7 | 439 | 529 | 529 | 428 | 378 | 429 | 474 | 418 | 463 |
| 8 | 474 | 544 | 544 | 465 | 423 | 467 | 584 | 438 | 482 |
| 9 | 499 | 536 | 536 | 478 | 446 | 482 | 482 | 459 | 464 |
| 10 | 519 | 596 | 596 | 486 | 488 | 497 | 488 | 479 | 478 |
| 11 | 538 |  |  | 505 | 532 | 517 | 533 | 500 | 520 |
| 12 | 561 |  |  | 531 | 554 | 533 |  | 517 | 504 |
| 13 | 578 |  |  | 580 | 605 | 634 |  | 536 | 486 |
| 14 | 610 |  |  | 513 |  | 494 |  | 553 | 422 |
| 15+ | 643 |  |  | 615 |  |  |  | 614 | 604 |
|  |  |  |  |  |  |  |  |  |  |
| AGE | 4.c | 5.a | 5.b | $6 . a$ | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 177 |  |  | 181 |  | 113 | 181 |  | 174 |
| 2 | 263 | 369 | 369 | 204 | 276 | 185 | 283 | 276 | 261 |
| 3 | 280 | 369 | 369 | 318 | 276 | 225 | 316 | 276 | 280 |
| 4 | 325 | 401 | 401 | 369 | 277 | 256 | 367 | 276 | 322 |
| 5 | 339 | 419 | 419 | 305 | 419 | 298 | 381 |  | 339 |
| 6 | 359 | 433 | 433 | 300 | 433 | 292 | 402 |  | 357 |
| 7 | 412 | 454 | 454 | 307 | 454 | 302 | 518 | 527 | 409 |
| 8 | 403 | 485 | 485 | 353 | 485 | 338 | 557 | 562 | 398 |
| 9 | 435 | 495 | 495 | 375 | 495 | 373 | 581 | 590 | 432 |
| 10 | 441 | 516 | 516 | 451 | 516 | 450 | 568 | 573 | 438 |
| 11 | 535 | 539 | 539 | 370 | 539 | 369 | 578 | 578 | 537 |
| 12 | 569 | 564 | 564 | 545 | 564 |  | 611 | 611 | 571 |
| 13 | 612 | 670 | 670 | 618 | 670 |  | 617 | 617 | 613 |
| 14 | 613 |  |  | 494 |  |  | 719 | 719 | 613 |
| 15+ | 602 |  |  |  |  |  |  |  | 602 |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017 (cont.).
Quarter 3

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  | 19 | 19 |  |
| 1 | 186 | 113 | 181 |  |  | 148 | 143 |  |
| 2 | 194 | 164 | 276 | 289 | 276 | 276 | 244 | 228 |
| 3 | 239 | 226 | 277 | 342 | 276 | 276 | 283 | 279 |
| 4 | 266 | 256 | 281 | 374 | 276 | 276 | 310 | 313 |
| 5 | 297 | 286 | 381 | 394 |  |  | 373 | 368 |
| 6 | 329 | 308 | 402 | 412 |  | 411 | 399 |  |
| 7 | 462 | 327 | 473 | 525 | 527 |  | 395 | 396 |
| 8 | 505 | 379 | 522 | 562 | 562 |  | 417 | 418 |
| 9 | 522 | 370 | 534 | 590 | 590 |  | 490 | 489 |
| 10 | 533 | 492 | 535 | 573 | 573 |  |  |  |
| 11 | 578 | 578 | 576 | 578 | 578 |  | 588 | 584 |
| 12 | 611 | 611 | 609 | 611 | 611 |  |  |  |
| 13 | 617 | 617 | 617 | 617 | 617 |  |  |  |
| 14 | 719 | 719 | 717 | 719 | 719 |  |  |  |
| $15+$ | 602 |  |  |  |  |  |  |  |


| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | 14.a | 14.b | All |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 60 |  |  | 57 | 72 |  |  | 22 |
| 1 | 134 | 137 | 200 | 111 | 123 | 431 | 431 | 311 |
| 2 | 230 | 269 | 232 | 300 | 223 | 369 | 369 | 331 |
| 3 | 268 | 282 | 287 | 372 | 273 | 422 | 422 | 348 |
| 4 | 298 | 316 | 337 | 448 | 314 | 481 | 479 | 377 |
| 5 | 364 | 357 | 360 | 582 | 346 | 532 | 523 | 410 |
| 6 | 401 | 407 | 375 |  | 373 | 577 | 516 | 426 |
| 7 | 402 | 451 | 412 |  | 386 | 644 | 477 | 442 |
| 8 | 420 | 450 | 418 |  | 413 |  | 501 | 476 |
| 9 |  |  | 429 |  |  |  | 510 | 497 |
| 10 | 464 | 453 |  |  |  |  | 532 | 517 |
| 11 |  |  |  |  |  |  | 554 | 537 |
| 12 | 514 | 530 |  |  |  | 581 | 560 |  |
| 13 |  |  |  |  |  |  | 578 |  |
| 14 |  |  |  |  |  |  | 611 |  |
| $15+$ |  |  |  |  |  |  | 642 |  |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017 (cont.).

## Quarter 4

| Age | 2.a | 2.a. 1 | 2.a. 2 | 3.a | 3.b | 3.c | 3.d | 4.a | 4.b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 231 | 318 | 318 | 184 | 177 | 177 | 191 | 198 | 200 |
| 2 | 307 | 316 | 316 | 284 | 250 | 272 | 290 | 271 | 261 |
| 3 | 311 | 303 | 303 | 318 | 300 | 309 | 326 | 305 | 286 |
| 4 | 351 | 385 | 385 | 370 | 389 | 368 | 376 | 367 | 357 |
| 5 | 381 | 423 | 423 | 383 | 384 | 384 | 389 | 382 | 381 |
| 6 | 401 | 436 | 436 | 406 | 377 | 397 | 420 | 405 | 412 |
| 7 | 420 | 446 | 446 | 432 | 396 | 429 | 441 | 426 | 436 |
| 8 | 440 | 401 | 401 | 475 | 448 | 467 | 497 | 458 | 443 |
| 9 | 462 | 489 | 489 | 486 | 450 | 482 | 488 | 472 | 451 |
| 10 | 487 | 580 | 580 | 490 | 488 | 497 | 484 | 483 | 461 |
| 11 | 504 | 612 | 612 | 509 | 533 | 517 | 508 | 506 | 521 |
| 12 | 519 | 542 | 542 | 542 |  | 533 | 554 | 532 | 545 |
| 13 | 538 |  |  | 611 |  | 634 | 605 | 571 | 586 |
| 14 | 554 |  |  | 496 |  | 494 |  | 528 | 589 |
| 15+ | 616 |  |  | 616 |  |  |  | 612 | 603 |
|  |  |  |  |  |  |  |  |  |  |
| AGE | 4.c | 5.a | 5.b | 6.a | 6.b | 7.a | 7.b | 7.c | 7.d |
| 0 |  |  |  |  |  |  |  |  |  |
| 1 | 173 |  | 170 | 186 |  | 125 | 181 |  | 179 |
| 2 | 267 | 370 | 243 | 253 |  | 156 | 283 |  | 260 |
| 3 | 278 | 341 | 292 | 293 |  | 222 | 318 |  | 279 |
| 4 | 320 | 391 | 314 | 385 |  | 269 | 369 |  | 321 |
| 5 | 335 | 414 | 355 | 382 |  | 336 | 381 |  | 340 |
| 6 | 354 | 433 | 351 | 362 |  | 378 | 402 |  | 357 |
| 7 | 409 | 442 | 375 | 422 |  | 405 | 525 |  | 407 |
| 8 | 396 | 473 | 420 | 481 |  | 447 | 561 |  | 398 |
| 9 | 433 | 492 | 444 | 489 |  | 448 | 588 |  | 434 |
| 10 | 439 | 508 | 488 | 524 |  | 483 | 572 |  | 442 |
| 11 | 537 | 525 | 533 | 575 |  | 516 | 578 |  | 537 |
| 12 | 571 | 575 |  | 608 |  | 533 | 611 |  | 571 |
| 13 | 613 | 686 |  | 617 |  | 634 | 617 |  | 613 |
| 14 | 613 |  |  | 719 |  | 494 | 719 |  | 613 |
| 15+ | 602 |  |  |  |  |  |  |  | 602 |

Table 8.5.2.1. NE Atlantic Mackerel. Mean weight (g) -at-age by area for 2017 (cont.).
Quarter 4

| Age | 7.e | 7.f | 7.g | 7.h | 7.j | 7.k | 8.a | 8.b |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 |  |  |  |  |  | 19 | 19 |  |
| 1 | 174 | 85 | 177 |  |  | 142 | 142 |  |
| 2 | 242 | 181 | 308 | 276 | 276 | 180 | 195 |  |
| 3 | 272 | 227 | 364 | 276 | 276 | 254 | 269 |  |
| 4 | 322 | 256 | 386 |  |  | 310 | 314 |  |
| 5 | 351 | 282 | 400 |  | 360 | 364 |  |  |
| 6 | 441 | 312 | 438 | 527 | 527 | 378 | 387 |  |
| 7 | 450 | 369 | 478 | 562 | 562 | 397 | 397 |  |
| 8 | 497 | 343 | 490 | 590 | 590 | 419 | 419 |  |
| 9 | 484 | 438 | 504 | 573 | 573 |  |  |  |
| 10 | 578 | 578 | 560 | 578 | 578 |  |  |  |
| 11 | 611 | 611 | 597 | 611 | 611 | 548 | 548 |  |
| 12 | 617 | 617 | 618 | 617 | 617 |  |  |  |
| 13 | 719 | 719 | 683 | 719 | 719 |  |  |  |
| 14 | 602 |  |  |  |  |  |  |  |

$15+\quad 602$

| Age | 8.c | 8.c.E | 8.d | 9.a | 9.a.N | $14 . \mathrm{a}$ | $14 . \mathrm{b}$ | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 46 |  |  | 49 | 87 |  |  | 21 |
| 1 | 141 | 163 |  | 119 | 129 |  |  | 199 |
| 2 | 194 | 244 |  | 260 | 173 |  |  | 267 |
| 3 | 237 | 269 |  | 376 | 254 |  |  | 305 |
| 4 | 260 | 302 |  | 481 | 305 |  |  | 365 |
| 5 | 363 | 360 |  | 439 | 349 |  |  | 382 |
| 6 | 383 | 388 |  | 450 | 364 |  |  | 404 |
| 7 | 389 | 403 |  | 413 | 385 |  |  | 424 |
| 8 | 415 | 418 |  | 413 | 413 |  |  | 453 |
| 9 |  |  |  |  |  |  |  | 469 |
| 10 | 477 | 490 |  |  |  |  |  | 486 |
| 11 |  |  |  |  |  |  |  | 507 |
| 12 |  | 530 |  |  |  |  |  | 527 |
| 13 |  |  |  |  |  |  |  | 560 |
| 14 |  |  |  |  |  |  |  | 556 |
| 15+ |  |  |  |  |  |  |  | 614 |

Table 8.6.1.1.1 International mackerel and horse mackerel egg survey: Periods and area assignments for vessels by week for the 2019 survey. Area assignments and dates are provisional.

|  |  | Area |  |  |  |  |  |  | Period |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| week | Starts | Portugal, <br>  <br> Galicia | Cantabrian Sea | Bay of Biscay | Celtic Sea | North west Ireland | West of Scotland | Northern Area |  |
| 3 | 13-Jan-19 | PO1 (DEPM) |  |  |  |  |  |  | 1 |
| 4 | 20-Jan-19 | PO1 (DEPM) |  |  |  |  |  |  | 1 |
| 5 | 27-Jan-19 | PO1 (DEPM) |  | IRL1 | IRL1 |  |  |  | 2 |
| 6 | 03-feb-19 | PO1 (DEPM) |  | IRL1 | IRL1 |  |  |  | 2 |
| 7 | 10-feb-19 | PO1 (DEPM) |  | IRL1 | IRL1 |  |  |  | 2 |
| 8 | 17-feb-19 | PO1 (DEPM) |  |  |  | SCO (IBTS) | SCO (IBTS) |  | 2 |
| 9 | 24-feb-19 |  |  |  |  | SCO (IBTS) | SCO (IBTS) |  | 2 |
| 10 | 03-mar-19 |  |  |  |  |  |  |  | 3 |
| 11 | 10-mar-19 |  | IEO1 |  |  | SCO2 | SCO2 |  | 3 |
| 12 | 17-mar-19 |  | IEO1 | AZTI1 | GER1 | SCO2 | SCO2 |  | 3 |
| 13 | 24-mar-19 |  | IEO1 | AZTI1 | GER1 | GER1 |  |  | 3 |
| 14 | 31-mar-19 |  | IEO1 | AZTI1 | GER1 | GER1 |  |  | 3 |
| 15 | 07-Apr-19 |  |  | IEO2 |  |  |  |  | 4 |
|  |  |  | IEO2 | AZTI1 | GER2 | GER2 |  |  |  |
| 16 | 14-Apr-19 |  | IEO2 | IEO2 | GER2 | GER2 | DEN | DEN | 4 |
| 17 | 21-Apr-19 |  | IEO2 | IEO2 | GER2 | DEN | DEN | DEN | 4 |
| 18 | 28-Apr -19 |  | IEO2 | IEO2 |  |  |  |  | 4 |
| 19 | 05-may-19 |  | AZTI2 (DEPM) | AZTI2 (DEPM) | NED1 | SCO3 | SCO3 | ICE | 5 |
| 20 | 12-may-19 |  | AZTI2 (DEPM) | AZTI2 (DEPM) | NED1 | SCO3 | SCO3 | ICE | 5 |
| 21 | 19-may-19 |  | AZTI2 (DEPM) | AZTI2 (DEPM) | NED1 | SCO3 | SCO3 | FAR | 5 |
| 22 | 26-may-19 |  | AZTI2 (DEPM) | AZTI2 (DEPM) |  |  |  | FAR | 5 |
| 23 | 02-jun-19 |  |  | NED2 | NED2 |  |  | FAR | 5 |
| 24 | 09-jun-19 |  |  |  |  |  |  |  | 6 |
|  |  |  |  | NED2 | NED2 | IRL2 | IRL2 | NOR |  |
| 25 | 16-jun-19 |  |  | NED2 | NED2 | IRL2 | IRL2 | NOR | 6 |
| 26 | 23-jun-19 |  |  |  |  | IRL2 | IRL2 | NOR | 6 |
| 27 | 30-jun-19 |  |  |  |  |  |  |  | 6 |
| 28 | 07-jul-19 |  |  |  | SCO3 | SCO4 | SCO4 |  | 7 |
| 29 | 14 -Jul-19 |  |  |  | SCO3 | SCO4 | SCO4 |  | 7 |
| 30 | 21-jul-19 |  |  |  | SCO3 | SCO4 | SCO4 |  | 7 |
| 31 | 28-jul-19 |  |  |  |  |  |  |  | 7 |

Table 8.6.3.1. Abundance index, mean weight-at-age, and biomass index for mackerel from the IESSNS in 2007 and from 2010 to 2018.

| AGE | 2007 |  |  | 2010 |  |  | 2011 |  |  | 2012 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number (billions) | $\begin{aligned} & \mathrm{W} \\ & (\mathrm{~g}) \end{aligned}$ | Biom. <br> t <br> (mil- <br> lion) | Num- <br> ber <br> (bil- <br> lions) | W <br> (g) | Biom. <br> t <br> (mil- <br> lion) | Number (billions) | $\begin{aligned} & \mathrm{W} \\ & (\mathrm{~g}) \end{aligned}$ | Biom. <br> t <br> (mil- <br> lion) | Num- <br> ber <br> (bil- <br> lions) | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Biom. <br> t <br> (mil- <br> lion) |
| 1 | 1.33 | 133 | 0.18 | 0.03 | 133 | 0 | 0.21 | 133 | 0.03 | 0.5 | 112 | 0.06 |
| 2 | 1.86 | 233 | 0.43 | 2.8 | 212 | 0.59 | 0.26 | 278 | 0.07 | 4.99 | 188 | 0.94 |
| 3 | 0.9 | 323 | 0.29 | 1.52 | 290 | 0.44 | 0.87 | 318 | 0.28 | 1.22 | 286 | 0.35 |
| 4 | 0.24 | 390 | 0.09 | 4.02 | 353 | 1.42 | 1.11 | 371 | 0.41 | 2.11 | 347 | 0.73 |
| 5 | 1 | 472 | 0.47 | 3.06 | 388 | 1.19 | 1.64 | 412 | 0.67 | 1.82 | 397 | 0.72 |
| 6 | 0.16 | 532 | 0.09 | 1.35 | 438 | 0.59 | 1.22 | 440 | 0.54 | 2.42 | 414 | 1 |
| 7 | 0.06 | 536 | 0.03 | 0.53 | 512 | 0.27 | 0.57 | 502 | 0.29 | 1.64 | 437 | 0.72 |
| 8 | 0.04 | 585 | 0.02 | 0.39 | 527 | 0.2 | 0.28 | 537 | 0.15 | 0.65 | 458 | 0.3 |
| 9 | 0.03 | 591 | 0.02 | 0.2 | 548 | 0.11 | 0.12 | 564 | 0.07 | 0.34 | 488 | 0.17 |
| 10 | 0.01 | 640 | 0.01 | 0.05 | 580 | 0.03 | 0.07 | 541 | 0.04 | 0.12 | 523 | 0.06 |
| 11 | 0.01 | 727 | 0.01 | 0.03 | 645 | 0.02 | 0.06 | 570 | 0.03 | 0.07 | 514 | 0.03 |
| 12 | 0 | 656 | 0 | 0.02 | 683 | 0.01 | 0.02 | 632 | 0.01 | 0.02 | 615 | 0.01 |
| 13 | 0.01 | 685 | 0.01 | 0.01 | 665 | 0.01 | 0.01 | 622 | 0.01 | 0.01 | 509 | 0 |
| 14+ | 0 | 671 | 0 | 0.01 | 596 | 0 | 0 | 612 | 0 | 0.01 | 677 | 0 |
| $\begin{aligned} & \text { TO- } \\ & \text { TAL } \end{aligned}$ | 5.65 | 512 | 1.64 | 13.99 | 469 | 4.89 | 6.42 | 467 | 2.69 | 15.91 | 426 | 5.09 |
|  | 2013 |  |  | 2014 |  |  | 2015 |  |  | 2016 |  |  |
| AGE | Num- <br> ber <br> (bil- <br> lions) | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Biom. <br> t <br> (mil- <br> lion) | Num- <br> ber <br> (bil- <br> lions) | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Biom. <br> t <br> (mil- <br> lion) | Number (billions) | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Biom. <br> t <br> (mil- <br> lion) | Number (billions) | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Biom. <br> t <br> (mil- <br> lion) |
| 1 | 0.06 | 96 | 0.01 | 0.01 | 228 | 0 | 1.2 | 128 | 0.15 | <0.01 | 95 | <0.01 |
| 2 | 7.78 | 184 | 1.43 | 0.58 | 275 | 0.16 | 0.83 | 290 | 0.24 | 4.98 | 231 | 1.15 |
| 3 | 8.99 | 259 | 2.32 | 7.8 | 288 | 2.24 | 2.41 | 333 | 0.8 | 1.37 | 324 | 0.45 |
| 4 | 2.14 | 326 | 0.7 | 5.14 | 335 | 1.72 | 5.77 | 342 | 1.97 | 2.64 | 360 | 0.95 |
| 5 | 2.91 | 374 | 1.09 | 2.61 | 402 | 1.05 | 4.56 | 386 | 1.76 | 5.24 | 371 | 1.95 |
| 6 | 2.87 | 399 | 1.15 | 2.62 | 433 | 1.14 | 1.94 | 449 | 0.87 | 4.37 | 394 | 1.72 |
| 7 | 2.68 | 428 | 1.15 | 2.67 | 459 | 1.23 | 1.83 | 463 | 0.85 | 1.89 | 440 | 0.83 |
| 8 | 1.27 | 445 | 0.56 | 1.69 | 477 | 0.8 | 1.04 | 479 | 0.5 | 1.66 | 458 | 0.76 |
| 9 | 0.45 | 486 | 0.22 | 0.74 | 488 | 0.36 | 0.62 | 488 | 0.3 | 1.11 | 479 | 0.53 |
| 10 | 0.19 | 523 | 0.1 | 0.36 | 533 | 0.19 | 0.32 | 505 | 0.16 | 0.75 | 488 | 0.37 |
| 11 | 0.16 | 499 | 0.08 | 0.09 | 603 | 0.05 | 0.08 | 559 | 0.04 | 0.45 | 494 | 0.22 |
| 12 | 0.04 | 547 | 0.02 | 0.05 | 544 | 0.03 | 0.07 | 568 | 0.04 | 0.2 | 523 | 0.1 |
| 13 | 0.01 | 677 | 0.01 | 0.02 | 537 | 0.01 | 0.04 | 583 | 0.02 | 0.07 | 511 | 0.04 |
| 14+ | 0.02 | 607 | 0.01 | 0 | 569 | 0 | 0.02 | 466 | 0.01 | 0.07 | 664 | 0.04 |
| TO- <br> TAL | 29.57 | 418 | 8.85 | 24.37 | 441 | 8.98 | 20.72 | 431 | 7.72 | 24.81 | 367 | 9.11 |

Table 8.6.3.1. Abundance index, mean weight-at-age, and biomass index for mackerel from the IESSNS in 2007 and from 2010 to 2018. Cont.

|  | 2017 |  |  | 2018 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | Num- <br> ber <br> (bil- <br> lions) | W <br> (g) | Biom. t <br> (mil- <br> lion) | Num- <br> ber <br> (bil- <br> lions) | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Biom. t (million) |
| 1 | 0.86 | 86 | 0.07 | 2.18 | 67 | 0.15 |
| 2 | 0.12 | 292 | 0.03 | 2.5 | 229 | 0.57 |
| 3 | 3.56 | 330 | 1.18 | 0.5 | 330 | 0.16 |
| 4 | 1.95 | 373 | 0.73 | 2.38 | 390 | 0.93 |
| 5 | 3.32 | 431 | 1.43 | 1.2 | 420 | 0.5 |
| 6 | 4.68 | 437 | 2.04 | 1.41 | 449 | 0.63 |
| 7 | 4.65 | 462 | 2.15 | 2.33 | 458 | 1.07 |
| 8 | 1.75 | 487 | 0.86 | 1.79 | 477 | 0.85 |
| 9 | 1.94 | 536 | 1.04 | 1.05 | 486 | 0.51 |
| 10 | 0.63 | 534 | 0.33 | 0.5 | 515 | 0.26 |
| 11 | 0.51 | 542 | 0.28 | 0.56 | 534 | 0.3 |
| 12 | 0.12 | 574 | 0.07 | 0.29 | 543 | 0.16 |
| 13 | 0.08 | 589 | 0.05 | 0.14 | 575 | 0.08 |
| 14+ | 0.04 | 626 | 0.03 | 0.09 | 643 | 0.05 |
| $\begin{aligned} & \text { TO- } \\ & \text { TAL } \end{aligned}$ | 24.22 | 425 | 10.29 | 16.92 | 368 | 6.22 |

Table 8.6.4.1. Numbers of RFID tagged and recaptured (by 31.08.2018) mackerel by tagging experiment. In the 2018 tagging survey off Ireland-Hebrides a proportion of the tagged mackerel were handled in the old way (marked * in the table), with manual jigging, and released directly to the sea at starboard side. This was to test whether differences in survival rates between the steel tag time series and the RFID tag time series is due to handling.

| Year | Period | Area | N-Released | N-Recaptured |
| :---: | :--- | :--- | ---: | ---: |
| 2011 | May-June | Ireland-Hebrides | 18645 | 133 |
| 2011 | Sep | Norwegian west coast | 31253 | 144 |
| 2012 | May-June | Ireland-Hebrides | 32137 | 276 |
| 2013 | May-June | Ireland-Hebrides | 22792 | 328 |
| 2014 | May-June | Ireland-Hebrides | 55185 | 885 |
| 2015 | May-June | Ireland-Hebrides | 43910 | 561 |
| 2015 | August | Iceland | 806 | 11 |
| 2016 | May-June | Ireland-Hebrides | 43959 | 537 |
| 2016 | August | Iceland | 4884 | 119 |
| 2017 | May-June | Ireland-Hebrides | 56082 | 286 |
| 2017 | August | Iceland | 3891 | 43 |
| 2018 | May-June | Ireland-Hebrides | 35336 | 13 |
| $2018 *$ | May-June | Ireland-Hebrides | 4661 | 1 |
| Total |  |  |  | 353541 |

Table 8.6.4.2. Numbers of recaptured mackerel with RFID tags by factory and recapture year.

| Factory | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | Total |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| DK01 Sæby | 0 | 0 | 8 | 11 | 0 | 0 | 0 | 19 |
| FO01 Vardin Pelagic | 0 | 0 | 15 | 37 | 23 | 13 | 0 | 88 |
| GB01 Denholm Coldstore | 0 | 0 | 0 | 10 | 10 | 28 | 25 | 73 |
| GB01 Denholm Factory | 0 | 0 | 25 | 64 | 79 | 119 | 31 | 318 |
| GB02 Lunar Freezing Peterhead | 0 | 0 | 33 | 51 | 60 | 42 | 20 | 206 |
| GB03 Lunar Freezing Fraserburgh | 0 | 0 | 0 | 9 | 16 | 7 | 5 | 37 |
| GB04 Pelagia Shetland | 0 | 0 | 25 | 130 | 162 | 157 | 53 | 527 |
| GB05 Northbay Pelagic | 0 | 0 | 0 | 0 | 0 | 0 | 23 | 23 |
| IC01 Vopnafjord | 0 | 0 | 24 | 61 | 81 | 73 | 37 | 276 |
| IC02 Neskaupstad | 0 | 0 | 0 | 19 | 93 | 58 | 23 | 193 |
| IC03 Höfn | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 2 |
| NO01 Pelagia Egersund Seafood | 12 | 25 | 19 | 7 | 1 | 0 | 53 | 117 |
| NO02 Skude Fryseri | 6 | 9 | 21 | 19 | 27 | 55 | 16 | 153 |
| NO03 Pelagia Austevoll | 1 | 1 | 7 | 5 | 1 | 0 | 3 | 18 |
| NO04 Pelagia Florø | 6 | 19 | 33 | 22 | 18 | 0 | 0 | 98 |
| NO05 Pelagia Måløy | 6 | 19 | 21 | 46 | 42 | 89 | 7 | 230 |
| NO06 Pelagia Selje | 19 | 35 | 38 | 77 | 59 | 102 | 24 | 354 |
| NO07 Pelagia Liavågen | 10 | 13 | 34 | 34 | 30 | 102 | 0 | 223 |
| NO08 Brødrene Sperre | 7 | 18 | 21 | 66 | 117 | 85 | 30 | 344 |
| NO09 Lofoten Viking | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 7 |
| NO14 Nils Sperre | 0 | 0 | 0 | 0 | 0 | 0 | 30 | 30 |
| NO16 Vikomar | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| All factories summed | 67 | 139 | 324 | 669 | 819 | 931 | 388 | 3337 |

Table 8.6.4.3. Abundance index in billions individuals ages $4-12$ per release years 2011-2016. The index is based on RFID tagging experiments and data from scanned catches and recaptures the two first years after a relase year (yearsout=1-2). The index is already scaled down to the $10 \%$ survival estimated by SAM.

|  |  |  |  | Age |  |  |  |  |  |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| 2011 | 3,236171 | 2,813887 | 2,017941 | 0,635952 | 0,322985 | 0,137462 | 0,036253 | 0,041593 | 0,064976 |
| 2012 | 3,484761 | 3,284821 | 2,956715 | 1,741262 | 0,572768 | 0,361678 | 0,128663 | 0,056303 | 0,036048 |
| 2013 | 1,974994 | 2,161885 | 1,956662 | 1,423692 | 0,709068 | 0,246734 | 0,141168 | 0,040904 | 0,009638 |
| 2014 | 3,206810 | 1,412106 | 1,636556 | 1,360554 | 0,890053 | 0,437290 | 0,191334 | 0,077791 | 0,037645 |
| 2015 | 2,696358 | 2,484074 | 1,350807 | 1,331608 | 1,065801 | 0,835234 | 0,400878 | 0,219851 | 0,061100 |
| 2016 | 0,890211 | 1,764716 | 1,618460 | 0,741481 | 0,727085 | 0,506326 | 0,355878 | 0,176446 | 0,048090 |

Table 8.6.5.2.1. Biomass, abundance, mean length and mean weight at age of mackerel from the Spanish spring acoustics surveys (PELACUS) from 2001 to 2018.

|  | 2001 |  |  |  | 2002 |  |  |  | 2003 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | Num- <br> ber <br> (mil- <br> lions) | $\begin{aligned} & \mathrm{L} \\ & (\mathrm{~cm}) \end{aligned}$ | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Bio- <br> mass <br> t <br> ('000) | Num- <br> ber <br> (mil- <br> lions) | $\begin{aligned} & \mathrm{L} \\ & (\mathrm{~cm}) \end{aligned}$ | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Bio- <br> mass <br> t <br> ('000) | Num- <br> ber <br> (mil- <br> lions) | $\begin{aligned} & \mathrm{L} \\ & (\mathrm{~cm}) \end{aligned}$ | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Bio- <br> mass <br> t <br> ('000) |
| 1 | 29.0 | 25.9 | 126.2 | 3.7 | 621.4 | 23.3 | 80.5 | 50.0 | 5678.6 | 23.1 | 81.6 | 463.2 |
| 2 | 47.6 | 31.0 | 213.7 | 10.2 | 94.8 | 32.0 | 221.9 | 21.0 | 324.5 | 28.9 | 165.1 | 53.6 |
| 3 | 184.3 | 33.7 | 277.3 | 51.1 | 378.1 | 34.3 | 277.1 | 104.8 | 109.0 | 33.5 | 261.3 | 28.5 |
| 4 | 386.6 | 36.1 | 340.3 | 131.6 | 706.8 | 35.8 | 317.9 | 224.7 | 229.0 | 35.0 | 299.7 | 68.6 |
| 5 | 382.1 | 37.5 | 383.0 | 146.4 | 1065.9 | 36.8 | 348.0 | 370.9 | 265.2 | 37.1 | 359.1 | 95.2 |
| 6 | 393.6 | 38.0 | 397.7 | 156.5 | 604.6 | 38.2 | 390.9 | 236.3 | 230.1 | 38.0 | 385.7 | 88.8 |
| 7 | 202.7 | 39.5 | 446.7 | 90.5 | 674.5 | 39.1 | 419.2 | 282.8 | 94.3 | 39.8 | 443.4 | 41.8 |
| 8 | 143.5 | 40.0 | 464.5 | 66.7 | 191.4 | 39.9 | 447.2 | 85.6 | 88.5 | 40.1 | 454.6 | 40.2 |
| 9 | 83.7 | 40.5 | 481.7 | 40.3 | 158.4 | 40.3 | 461.4 | 73.1 | 19.6 | 41.5 | 505.1 | 9.9 |
| 10 | 17.0 | 40.2 | 469.3 | 8.0 | 100.2 | 41.0 | 490.2 | 49.1 | 10.0 | 41.9 | 519.9 | 5.2 |
| 11 | 26.3 | 42.1 | 541.4 | 14.2 | 54.0 | 41.4 | 504.0 | 27.2 | 14.0 | 42.6 | 549.6 | 7.7 |
| 12 | 12.3 | 41.9 | 533.8 | 6.5 | 12.4 | 43.5 | 586.7 | 7.3 | 3.8 | 41.5 | 503.1 | 1.9 |
| 13 | 1.9 | 41.5 | 517.1 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3.7 | 43.1 | 566.9 | 2.1 |
| 14 | 6.1 | 43.5 | 596.5 | 3.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 15+ | 9.4 | 42.8 | 568.1 | 5.3 | 2.9 | 45.5 | 676.9 | 2.0 | 2.0 | 43.3 | 578.1 | 1.2 |
| $\begin{aligned} & \text { TO- } \\ & \text { TAL } \end{aligned}$ | 1926.2 | 37.3 | 381.9 | 735.6 | 4665.3 | 35.5 | 329.0 | 1534.8 | 7072.1 | 25.5 | 128.4 | 907.8 |
|  | 2004 |  |  |  | 2005 |  |  |  | 2006 |  |  |  |
| 1 | 195.2 | 25.0 | 114.6 | 22.4 | 43.4 | 24.8 | 112.1 | 4.6 | 83.7 | 20.8 | 58.5 | 4.9 |
| 2 | 952.4 | 28.3 | 164.5 | 156.6 | 106.5 | 29.2 | 181.8 | 19.0 | 9.3 | 29.7 | 177.2 | 1.7 |
| 3 | 599.3 | 32.8 | 258.1 | 154.7 | 229.1 | 32.3 | 245.4 | 56.1 | 57.3 | 31.9 | 223.1 | 12.8 |
| 4 | 227.5 | 37.5 | 377.8 | 86.0 | 259.6 | 36.5 | 349.4 | 92.4 | 230.7 | 33.5 | 262.7 | 60.6 |
| 5 | 425.6 | 38.1 | 395.5 | 168.3 | 82.6 | 38.3 | 403.4 | 34.2 | 104.7 | 36.7 | 345.0 | 36.1 |
| 6 | 336.7 | 39.1 | 428.4 | 144.2 | 163.8 | 38.8 | 417.6 | 70.4 | 34.2 | 38.5 | 398.1 | 13.6 |
| 7 | 181.5 | 40.1 | 461.7 | 83.8 | 114.9 | 39.5 | 438.4 | 52.0 | 22.2 | 39.2 | 420.5 | 9.3 |
| 8 | 106.1 | 40.8 | 483.2 | 51.3 | 63.8 | 39.8 | 451.7 | 29.8 | 7.6 | 40.9 | 483.3 | 3.6 |
| 9 | 76.5 | 41.0 | 492.5 | 37.7 | 33.6 | 41.0 | 493.9 | 17.2 | 2.0 | 41.9 | 513.6 | 1.0 |
| 10 | 31.1 | 42.3 | 538.0 | 16.7 | 15.3 | 42.3 | 535.4 | 8.5 | 3.4 | 41.3 | 495.1 | 1.7 |
| 11 | 18.9 | 42.2 | 533.9 | 10.1 | 13.7 | 41.8 | 518.8 | 7.4 | 1.4 | 42.7 | 545.7 | 0.8 |
| 12 | 13.5 | 43.3 | 573.8 | 7.7 | 6.6 | 42.0 | 526.6 | 3.6 | 0.5 | 42.8 | 551.1 | 0.3 |
| 13 | 3.2 | 43.9 | 599.8 | 1.9 | 11.3 | 42.5 | 544.1 | 6.4 | 0.1 | 43.8 | 590.7 | 0.1 |
| 14 | 0.0 | 0.0 | 0.0 | 0.0 | 5.1 | 43.8 | 592.6 | 3.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| 15+ | 5.9 | 46.4 | 710.5 | 4.2 | 7.3 | 43.7 | 594.9 | 4.6 | 0.0 | 44.5 | 621.0 | 0.0 |
| TO- <br> TAL | 3173.2 | 33.8 | 298.0 | 945.6 | 1156.6 | 35.9 | 346.7 | 409.5 | 557.3 | 32.7 | 263.0 | 146.6 |

Table 8.6.5.2.1. Biomass, abundance, mean length and mean weight at age of mackerel from the Spanish spring acoustics surveys (PELACUS) from 2001 to 2018 (cont.).

|  | 2007 |  |  |  | 2008 |  |  |  | 2009 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | Num- <br> ber <br> (mil- <br> lions) | $\begin{aligned} & \mathrm{L} \\ & (\mathrm{~cm}) \end{aligned}$ | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Biomass t ('000) | Num- <br> ber <br> (mil- <br> lions) | L (cm) | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Bio- <br> mass <br> t <br> ('000) | Num- <br> ber <br> (mil- <br> lions) | L (cm) | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Bio- <br> mass <br> t <br> ('000) |
| 1 | 182.2 | 21.5 | 64.1 | 11.7 | 407.1 | 24.4 | 100.4 | 40.9 | 7.5 | 24.3 | 98.5 | 0.7 |
| 2 | 34.6 | 25.6 | 110.5 | 3.8 | 100.5 | 27.1 | 135.2 | 13.6 | 65.1 | 29.3 | 176.1 | 11.5 |
| 3 | 22.1 | 33.4 | 254.5 | 5.6 | 327.4 | 29.8 | 180.7 | 59.1 | 148.4 | 30.0 | 189.4 | 28.1 |
| 4 | 129.6 | 34.9 | 291.7 | 37.8 | 125.8 | 33.5 | 261.9 | 32.9 | 201.7 | 32.5 | 248.1 | 50.0 |
| 5 | 189.4 | 36.1 | 324.0 | 61.4 | 233.6 | 36.2 | 328.2 | 76.5 | 86.8 | 35.0 | 314.3 | 27.3 |
| 6 | 117.5 | 38.1 | 379.7 | 44.6 | 277.5 | 36.3 | 328.5 | 91.0 | 148.8 | 36.9 | 370.0 | 55.0 |
| 7 | 31.9 | 39.8 | 435.9 | 13.9 | 131.0 | 37.9 | 374.1 | 48.9 | 180.8 | 37.7 | 394.7 | 71.3 |
| 8 | 20.5 | 39.7 | 431.5 | 8.8 | 25.2 | 39.5 | 423.4 | 10.6 | 93.0 | 39.5 | 454.8 | 42.2 |
| 9 | 4.8 | 41.2 | 484.0 | 2.3 | 20.1 | 39.5 | 422.7 | 8.5 | 32.6 | 40.2 | 484.7 | 15.7 |
| 10 | 6.1 | 40.7 | 464.7 | 2.8 | 20.5 | 40.2 | 443.6 | 9.0 | 14.9 | 40.7 | 500.8 | 7.5 |
| 11 | 1.5 | 41.4 | 490.3 | 0.8 | 9.2 | 41.1 | 474.8 | 4.4 | 4.6 | 41.6 | 537.0 | 2.4 |
| 12 | 4.7 | 44.5 | 608.6 | 2.8 | 7.3 | 41.8 | 500.0 | 3.6 | 3.5 | 42.2 | 561.9 | 2.0 |
| 13 | 0.7 | 43.5 | 567.6 | 0.4 | 2.4 | 43.4 | 561.4 | 1.3 | 4.1 | 42.4 | 569.2 | 2.3 |
| 14 | 2.6 | 44.0 | 591.5 | 1.5 | 1.1 | 44.6 | 607.1 | 0.7 | 0.0 | 0.0 | 0.0 | 0.0 |
| 15+ | 0.7 | 46.5 | 697.9 | 0.5 | 0.4 | 46.5 | 690.3 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 |
| $\begin{aligned} & \text { TO- } \\ & \text { TAL } \end{aligned}$ | 748.9 | 32.5 | 265.4 | 198.8 | 1689.2 | 31.7 | 238.0 | 401.4 | 991.8 | 34.8 | 319.0 | 316.2 |
|  | 2010 |  |  |  | 2011 |  |  |  | 2012 |  |  |  |
| 1 | 431.8 | 23.6 | 89.2 | 38.6 | 1936.9 | 22.5 | 77.4 | 149.3 | 698.05 | 22.07 | 74.36 | 51.83 |
| 2 | 72.7 | 30.6 | 194.8 | 14.2 | 29.7 | 30.5 | 201.3 | 6.0 | 16.7 | 27.71 | 150.62 | 2.5 |
| 3 | 189.6 | 31.5 | 214.9 | 40.9 | 63.1 | 32.3 | 239.2 | 15.1 | 11.18 | 33.27 | 265.58 | 2.98 |
| 4 | 662.7 | 33.6 | 262.3 | 174.1 | 90.6 | 33.7 | 273.6 | 24.7 | 32.34 | 34.63 | 299.04 | 9.69 |
| 5 | 873.3 | 35.0 | 296.3 | 258.8 | 154.8 | 35.0 | 308.5 | 47.6 | 60.04 | 35.62 | 325.28 | 19.53 |
| 6 | 306.6 | 36.8 | 346.3 | 106.1 | 144.1 | 36.1 | 340.6 | 49.0 | 147.09 | 36.58 | 353.17 | 51.84 |
| 7 | 388.9 | 38.1 | 385.6 | 149.8 | 57.7 | 38.2 | 406.2 | 23.4 | 121.31 | 37.66 | 386.73 | 46.77 |
| 8 | 239.2 | 38.2 | 388.3 | 92.8 | 54.2 | 39.5 | 446.9 | 24.1 | 61.9 | 39.43 | 445.95 | 27.53 |
| 9 | 113.9 | 39.5 | 427.5 | 48.6 | 31.2 | 39.6 | 451.5 | 14.0 | 32.39 | 40.12 | 470.22 | 15.19 |
| 10 | 26.4 | 40.8 | 470.2 | 12.4 | 10.3 | 41.0 | 503.5 | 5.2 | 19.11 | 40.54 | 485.42 | 9.26 |
| 11 | 16.5 | 40.9 | 475.8 | 7.8 | 4.7 | 41.0 | 503.1 | 2.4 | 8.07 | 40.66 | 489.56 | 3.94 |
| 12 | 10.3 | 41.4 | 492.4 | 5.0 | 3.1 | 41.8 | 533.3 | 1.6 | 2.78 | 41.94 | 538.24 | 1.49 |
| 13 | 7.5 | 41.9 | 509.7 | 3.8 | 2.4 | 41.6 | 527.1 | 1.2 | 1.36 | 42.38 | 555.37 | 0.75 |
| 14 | 5.3 | 42.4 | 530.5 | 2.8 | 0.0 | 0.0 | 0.0 | 0.0 | 1.36 | 42.38 | 555.37 | 0.75 |
| 15+ | 3.0 | 43.1 | 557.7 | 1.7 | 0.0 | 0.0 | 0.0 | 0.0 | 1.19 | 44.53 | 649.03 | 0.78 |
| $\begin{aligned} & \text { TO- } \\ & \text { TAL } \end{aligned}$ | 3347.8 | 34.0 | 286.0 | 957.5 | 2582.9 | 25.8 | 141.2 | 363.7 | 1214.88 | 28.46 | 201.91 | 244.81 |

Table 8.6.5.2.1. Biomass, abundance, mean length and mean weight at age of mackerel from the Spanish spring acoustics surveys (PELACUS) from 2001 to 2018 (cont.).

|  | 2013 |  |  |  | 2014 |  |  |  | 2015 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | Num- <br> ber <br> (mil- <br> lions) | $\begin{aligned} & \mathrm{L} \\ & (\mathrm{~cm}) \end{aligned}$ | $\begin{aligned} & \mathrm{W} \\ & (\mathrm{~g}) \end{aligned}$ | Bio- <br> mass <br> t <br> ('000) | Num- <br> ber <br> (mil- <br> lions) | $\begin{aligned} & \mathrm{L} \\ & (\mathrm{~cm}) \end{aligned}$ | $\begin{aligned} & \text { W } \\ & (\mathrm{g}) \end{aligned}$ | Bio- <br> mass <br> t <br> ('000) | Num- <br> ber <br> (mil- <br> lions) | $\begin{aligned} & \mathrm{L} \\ & (\mathrm{~cm}) \end{aligned}$ | $\begin{aligned} & \mathrm{W} \\ & (\mathrm{~g}) \end{aligned}$ | Bio- <br> mass <br> t <br> ('000) |
| 1 | 99 | 24.5 | 93.0 | 9 | 68.1 | 22.5 | 71.5 | 5.1 | 101.38 | 22.34 | 69.55 | 7.50 |
| 2 | 653 | 26.5 | 119.1 | 81 | 42.8 | 32.0 | 217.4 | 9.1 | 11.91 | 31.88 | 214.66 | 2.60 |
| 3 | 123 | 28.6 | 152.4 | 20 | 157.4 | 32.3 | 223.7 | 34.6 | 43.16 | 32.69 | 232.42 | 10.20 |
| 4 | 114 | 34.2 | 267.6 | 31 | 340.4 | 33.3 | 245.5 | 81.9 | 112.36 | 34.05 | 264.52 | 29.81 |
| 5 | 228 | 35.3 | 296.0 | 68 | 675.8 | 34.5 | 275.3 | 181.7 | 299.50 | 35.09 | 290.94 | 86.92 |
| 6 | 235 | 36.2 | 322.3 | 76 | 581.1 | 36.1 | 318.0 | 179.5 | 348.66 | 36.40 | 326.84 | 112.95 |
| 7 | 178 | 36.7 | 335.3 | 60 | 502.4 | 36.6 | 333.9 | 163.0 | 344.06 | 37.03 | 345.17 | 117.63 |
| 8 | 64 | 37.6 | 361.4 | 23 | 246.9 | 36.7 | 335.2 | 80.4 | 164.59 | 37.02 | 344.84 | 56.24 |
| 9 | 11 | 38.1 | 378.2 | 4 | 84.5 | 38.2 | 381.8 | 31.3 | 71.17 | 38.37 | 386.31 | 27.15 |
| 10 | 8 | 40.0 | 439.4 | 4 | 33.1 | 39.2 | 414.3 | 13.3 | 29.50 | 39.17 | 412.51 | 12.00 |
| 11 | 3 | 40.8 | 470.1 | 1 | 34.7 | 39.4 | 420.9 | 14.2 | 29.95 | 39.24 | 414.69 | 12.25 |
| 12 | 2 | 41.2 | 490.3 | 1 | 34.7 | 39.4 | 420.9 | 14.2 | 29.95 | 39.24 | 414.69 | 12.25 |
| 13 |  |  |  |  |  |  |  |  |  |  |  | 0 |
| 14 |  |  |  |  |  |  |  |  |  |  |  | 0 |
| 15+ |  |  |  |  |  |  |  |  |  |  |  | 0 |
| $\begin{aligned} & \text { TO- } \\ & \text { TAL } \end{aligned}$ | 1718 | 31.2 | 200.2 | 379 | 2802.0 | 35.1 | 291.0 | 808.4 | 1586.20 | 35.40 | 299.24 | 487.49 |
|  | 2016 |  |  |  | 2017 |  |  |  | 2018 |  |  |  |
| 1 | 12.61 | 22.4 | 74.0 | 1.0 | 170.5 | 21.9 | 67.2 | 12.4 |  | 22.72 | 81.99 | 5.3 |
| 2 | 73.54 | 28.0 | 144.1 | 11.2 | 12.4 | 27.8 | 141.3 | 1.9 |  | 27.46 | 142.93 | 5.1 |
| 3 | 26.62 | 30.9 | 193.1 | 5.3 | 91.4 | 62.8 | 234.2 | 22.6 |  | 33.56 | 256.69 | 10.1 |
| 4 | 54.98 | 34.5 | 268.2 | 14.8 | 115.6 | 64.8 | 283.1 | 34.5 |  | 35.73 | 309.38 | 30.9 |
| 5 | 230.22 | 35.7 | 297.7 | 68.9 | 438.3 | 65.4 | 298.2 | 137.2 |  | 35.99 | 315.99 | 124.3 |
| 6 | 406.48 | 36.4 | 315.3 | 128.9 | 421.2 | 36.1 | 316.4 | 139.9 |  | 36.52 | 329.78 | 143.6 |
| 7 | 318.08 | 37.3 | 337.3 | 107.8 | 278.3 | 37.1 | 344.8 | 100.7 |  | 37.33 | 351.83 | 116.2 |
| 8 | 271.41 | 37.8 | 353.4 | 96.2 | 128.7 | 38.1 | 374.3 | 50.4 |  | 38.04 | 371.91 | 58.1 |
| 9 | 102.70 | 38.3 | 365.1 | 37.6 | 84.4 | 38.2 | 377.0 | 33.2 |  | 38.12 | 374.13 | 41.8 |
| 10 | 50.36 | 38.4 | 367.8 | 18.6 | 21.8 | 38.4 | 384.1 | 8.7 |  | 38.30 | 379.46 | 10.8 |
| 11 | 13.83 | 38.9 | 383.8 | 5.3 | 11.8 | 40.1 | 439.1 | 5.4 |  | 40.10 | 434.16 | 7.0 |
| 12 | 5.31 | 39.4 | 398.6 | 2.1 | 2.7 | 39.5 | 418.0 | 1.2 |  | 41.64 | 484.65 | 3.4 |
| 13 |  | - | - | - |  |  |  |  |  |  |  |  |
| 14 | - | - |  | - |  |  |  |  |  |  |  |  |
| 15+ | '- | - | - | - |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { TO- } \\ & \text { TAL } \end{aligned}$ | 1566.14 | 36.3 | 311.7 | 497.7 | 1777.0 | 34.7 | 280.4 | 548.2 |  | 36.10 | 318.83 | 556.53 |

Table 8.6.5.2.2. Mackerel abundance and biomass by ICES sub-divisions from Spanish spring acoustic surveys (PELACUS) from 2001 to 2018.

|  | ICES 9.A-N |  | ICES 8.c-W |  | 8.C-EW |  | 8.C-EE |  | TOTAL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abund. $\left(10^{9}\right)$ | Biomass $(\mathrm{kt})$ | Abund. $\left(10^{9}\right)$ | Biomass <br> (kt) | Abund. $\left(10^{9}\right)$ | Biomass (kt) | Abund. $\left(10^{9}\right)$ | Biomass $(\mathrm{kt})$ | Abund. $\left(10^{9}\right)$ | Biomass <br> (kt) |
| 2001 | 0.02 | 7.4 | 0.31 | 120.1 | 1.23 | 489.1 | 0.36 | 119.1 | 1.93 | 735.7 |
| 2002 | 0.00 | 0.0 | 0.82 | 333.7 | 3.80 | 1191.1 | 0.04 | 10.0 | 4.67 | 1534.8 |
| 2003 | 4.58 | 376.6 | 1.07 | 184.4 | 0.88 | 202.5 | 0.54 | 144.3 | 7.14 | 907.8 |
| 2004 | 0.61 | 118.6 | 1.03 | 304.3 | 1.50 | 515.7 | 0.03 | 7.0 | 3.17 | 945.6 |
| 2005 | 0.16 | 45.6 | 0.23 | 13.0 | 0.60 | 228.6 | 0.16 | 32.3 | 1.06 | 409.5 |
| 2006 | 0.01 | 0.7 | 0.39 | 100.5 | 0.15 | 41.5 | 0.02 | 4.0 | 0.56 | 146.6 |
| 2007 | 0.16 | 11.2 | 0.22 | 77.4 | 0.36 | 108.4 | 0.01 | 1.8 | 0.75 | 198.8 |
| 2008 | 0.16 | 21.4 | 0.38 | 109.0 | 0.84 | 235.0 | 0.05 | 4.2 | 1.42 | 369.7 |
| 2009 | 0.06 | 11.8 | 0.04 | 10.1 | 0.57 | 220.2 | 0.33 | 74.1 | 0.99 | 316.2 |
| 2010 | 0.38 | 34.2 | 0.88 | 293.7 | 2.09 | 628.6 | 0.00 | 1.0 | 3.35 | 957.5 |
| 2011 | 1.42 | 109.2 | 0.51 | 39.4 | 0.65 | 212.4 | 0.01 | 2.7 | 2.58 | 363.7 |
| 2012 | 0.61 | 45.03 | 0.02 | 1.3 | 0.57 | 190.7 | 0.02 | 7.8 | 1.21 | 244.8 |
| 2013 | 0.00 | 00.00 | 0.46 | 58.0 | 1.06 | 270.9 | 0.19 | 49.7 | 1.72 | 378.6 |
| 2014* | 0.02 | 2.4 | 0.03 | 3.0 |  |  | 2.75 | 803 | 2.80 | 808.4 |
| 2015* | 0.21 | 73.6 | 0.3 | 7.4 |  |  | 1.36 | 410 | 1.57 | 483.3 |
| 2016* | 0.00 | 0.2 | 0.09 | 13.7 |  |  | 1.48 | 484 | 1.57 | 498 |
| 2017* | . 17 | 14.7 | . 36 | 119.0 |  |  | 1.25 | 415 | 1.78 | 548.7 |
| 2018* | 0.10 | 27.8 | 0.01 | 031 |  |  | 1.55 | $528 *$ | 1.64 | 556.5 |

* Without split between 8.c-EW and 8.c-EE.

Table 8.7.1.1. NE Atlantic mackerel. Input data and parameters and the model configurations for the assessment.

| Input data types and characteristics: |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name |  | Year range |  | Age range | Variable f year | m year to |
| Catch in tonnes |  | 1980-2017 |  |  | Yes |  |
| Catch-at-age in numbers |  | 1980-2017 |  | 0-12+ | Yes |  |
| Weight-at-age in the commercial catch |  | 1980-2017 |  | 0-12+ | Yes |  |
| Weight-at-age of the spawning stock at spawning time. |  | 1980-2018 |  | 0-12+ | Yes |  |
| Proportion of natural mortality before spawning |  | 1980-2018 |  | 0-12+ | Yes |  |
| Proportion of fishing mortality before spawning |  | 1980-2018 |  | 0-12+ | Yes |  |
| Proportion mature-at-age |  | 1980-2018 |  | 0-12+ | Yes |  |
| Natural mortality |  | 1980-2018 |  | 0-12+ | No, fixed | t 0.15 |
| Tuning data: |  |  |  |  |  |  |
| Type | Name |  | Year range |  |  | Age range |
| Survey (SSB) | ICES Triennial Mackerel and Horse Mackerel Egg Survey |  | 1992, 1995, 1998, 2001, 2004, 2007, 2010, 2013,2016. |  |  | Not applicable (gives SSB) |
| Survey (abundance index) | IBTS Recruitment index (log transformed) |  | 1998-2015 |  |  | Age 0 |
| Survey (abundance index) | International Ecosystem Summer Survey in the Nordic Seas (IESSNS) |  | 2010, 2012-2018 |  |  | Ages 3-11 |
| Tagging/recapture | Norwegian tagging program |  | Steal tags : 1980 (release year)-2006 (recapture years) RFID tags : 2011 (release year) 2017 (recapture year) |  |  | Ages 2 and older (age at release) |

SAM parameter configuration :

| Setting | Value | Description |
| :---: | :---: | :---: |
| Coupling of fishing mortality states | 1/2/3/4/5/6/7/8/8/8/8/8/8 | Different $F$ states for ages 0 to 6, one same F state for ages 7 and older |
| Correlated random walks for the fishing mortalities | 0 | F random walk of different ages are independent |
| Coupling of catchability parameters | 0/0/0/0/0/0/0/0/0/0/0/0/0 <br> 1/0/0/0/0/0/0/0/0/0/0/0/0 <br> 2/0/0/0/0/0/0/0/0/0/0/0/0 <br> 0/0/0/3/4/5/6/7/8/9/10/10/0 | No catchability parameter for the catches <br> One catchability parameter estimated for the egg <br> One catchability parameter estimated for the recruitment index <br> One catchability parameter for each age group estimated for the IESSNS (age 3 to11) |
| Power law model | 0 | No power law model used for any |


| Coupling of fishing mortality random walk variances |  | of the surveys |
| :---: | :---: | :---: |
|  | 1/1/1/1/1/1/1/1/1/1/1/1/1 | Same variance used for the F random walk of all ages |
| Coupling of log abundance random walk variances | 1/2/2/2/2/2/2/2/2/2/2/2/2 | Same variance used for the log abundance random walk of all ages except for the recruits (age 0) |
| Coupling of the observation variances | 1/1/1/1/1/1/1/1/1/1/1/1/1 <br> 0/0/0/0/0/0/0/0/0/0/0/0/0 <br> 2/0/0/0/0/0/0/0/0/0/0/0/0 <br> 0/0/0/3/4/4/4/4/4/4/4/4/0 | Same observation variance for all ages in the catches <br> One observation variance for the egg survey <br> One observation variance for the recruitment index <br> 2 observation variances for the IESSNS (age 3 and ages 4 and older) |
| Stock recruitment model | 0 | No stock-recruiment model |
| Correlation structure | "ID", "ID", "ID", "AR" | Auto-regressive correlation structure for the IESSNS index, independent observations assumed for the other data sources |

Table 8.7.1.2. NE Atlantic Mackerel. CATCH IN NUMBER

|  | year |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 0 | 33101 | 56682 | 11180 | 7333 | 287287 | 81799 | 49983 | 7403 | 57644 | 65400 | 24246 | 10007 |
| 1 | 411327 | 276229 | 213936 | 47914 | 31901 | 268960 | 58126 | 40126 | 152656 | 64263 | 140534 | 58459 |
| 2 | 393025 | 502365 | 432867 | 668909 | 86064 | 20893 | 424563 | 156670 | 137635 | 312739 | 209848 | 212521 |
| 3 | 64549 | 231814 | 472457 | 433744 | 682491 | 58346 | 38387 | 663378 | 190403 | 207689 | 410751 | 206421 |
| 4 | 328206 | 32814 | 184581 | 373262 | 387582 | 445357 | 76545 | 56680 | 538394 | 167588 | 208146 | 375451 |
| 5 | 254172 | 184867 | 26544 | 126533 | 251503 | 252217 | 364119 | 89003 | 72914 | 362469 | 156742 | 188623 |
| 6 | 142978 | 173349 | 138970 | 20175 | 98063 | 165219 | 208021 | 244570 | 87323 | 48696 | 254015 | 129145 |
| 7 | 145385 | 116328 | 112476 | 90151 | 22086 | 62363 | 126174 | 150588 | 201021 | 58116 | 42549 | 197888 |
| 8 | 54778 | 125548 | 89672 | 72031 | 61813 | 19562 | 42569 | 85863 | 122496 | 111251 | 49698 | 51077 |
| 9 | 130771 | 41186 | 88726 | 48668 | 47925 | 47560 | 13533 | 34795 | 55913 | 68240 | 85447 | 43415 |
| 10 | 39920 | 146186 | 27552 | 49252 | 37482 | 37607 | 32786 | 19658 | 20710 | 32228 | 33041 | 70839 |
| 11 | 56210 | 31639 | 91743 | 19745 | 30105 | 26965 | 22971 | 25747 | 13178 | 13904 | 16587 | 29743 |
| 12 | 104927 | 199615 | 156121 | 132040 | 69183 | 97652 | 81153 | 63146 | 57494 | 35814 | 27905 | 52986 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| 0 | 43447 | 19354 | 25368 | 14759 | 37956 | 36012 | 61127 | 67003 | 36345 | 26034 | 70409 | 14744 |
| 1 | 83583 | 128144 | 147315 | 81529 | 119852 | 144390 | 99352 | 73597 | 102407 | 40315 | 222577 | 187997 |
| 2 | 156292 | 210319 | 221489 | 340898 | 168882 | 186481 | 229767 | 132994 | 142898 | 158943 | 70041 | 275661 |
| 3 | 356209 | 266677 | 306979 | 340215 | 333365 | 238426 | 264566 | 223639 | 275376 | 234186 | 367902 | 91075 |
| 4 | 266591 | 398240 | 267420 | 275031 | 279182 | 378881 | 323186 | 261778 | 390858 | 297206 | 350163 | 295777 |
| 5 | 306143 | 244285 | 301346 | 186855 | 177667 | 246781 | 361945 | 281041 | 295516 | 309937 | 262716 | 235052 |
| 6 | 156070 | 255472 | 184925 | 197856 | 96303 | 135059 | 207619 | 244212 | 241550 | 231804 | 237066 | 183036 |
| 7 | 113899 | 149932 | 189847 | 142342 | 119831 | 84378 | 118388 | 159019 | 175608 | 195250 | 151320 | 133595 |
| 8 | 138458 | 97746 | 106108 | 113413 | 55812 | 66504 | 72745 | 86739 | 106291 | 120241 | 118870 | 94168 |
| 9 | 51208 | 121400 | 80054 | 69191 | 59801 | 39450 | 47353 | 50613 | 52394 | 72205 | 79945 | 75701 |
| 10 | 36612 | 38794 | 57622 | 42441 | 25803 | 26735 | 24386 | 30363 | 31280 | 42529 | 43789 | 45951 |
| 11 | 40956 | 29067 | 20407 | 37960 | 18353 | 13950 | 16551 | 17048 | 18918 | 20546 | 21611 | 25797 |
| 12 | 68205 | 68217 | 57551 | 39753 | 30648 | 24974 | 22932 | 32446 | 34202 | 40706 | 40280 | 30890 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| 0 | 11553 | 12426 | 75651 | 19302 | 25886 | 17615 | 23453 | 30429 | 23872 | 11325 | 62100 | 6732 |
| 1 | 31421 | 46840 | 149425 | 88439 | 59899 | 36514 | 78605 | 62708 | 66196 | 47020 | 43173 | 104019 |
| 2 | 453133 | 135648 | 173646 | 190857 | 167748 | 113574 | 137101 | 115346 | 200167 | 235411 | 137788 | 124411 |
| 3 | 529753 | 668588 | 159455 | 220575 | 399086 | 455113 | 303928 | 322725 | 214043 | 399751 | 669949 | 248852 |
| 4 | 147973 | 293579 | 470063 | 215655 | 284660 | 616963 | 739221 | 469953 | 415884 | 370551 | 829399 | 579835 |
| 5 | 258177 | 120538 | 195594 | 455131 | 260314 | 319465 | 611729 | 654395 | 456404 | 442597 | 564508 | 646894 |
| 6 | 145899 | 121477 | 97061 | 203492 | 255675 | 224848 | 284788 | 488713 | 511270 | 429324 | 549985 | 450344 |
| 7 | 89856 | 63612 | 73510 | 77859 | 124382 | 194326 | 143039 | 244210 | 323835 | 336701 | 503300 | 415107 |
| 8 | 65669 | 38763 | 33399 | 59652 | 57297 | 73171 | 102072 | 113012 | 142948 | 188910 | 339538 | 355997 |
| 9 | 40443 | 23947 | 18961 | 30494 | 32343 | 29738 | 45841 | 53363 | 69551 | 112765 | 141344 | 205691 |
| 10 | 35654 | 18612 | 13987 | 16039 | 19482 | 14989 | 21222 | 25046 | 30619 | 45938 | 63614 | 107685 |
| 11 | 16430 | 7955 | 8334 | 11416 | 6798 | 7470 | 6255 | 12311 | 11603 | 18928 | 21294 | 26939 |
| 12 | 19509 | 10669 | 10186 | 12801 | 9581 | 5003 | 8523 | 10775 | 11678 | 17857 | 13136 | 22700 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 2016 | 2017 |  |  |  |  |  |  |  |  |  |  |
| 0 | 716 | 28306 |  |  |  |  |  |  |  |  |  |  |
| 1 | 45199 | 43458 |  |  |  |  |  |  |  |  |  |  |
| 2 | 203753 | 87739 |  |  |  |  |  |  |  |  |  |  |
| 3 | 257293 | 458301 |  |  |  |  |  |  |  |  |  |  |
| 4 | 424843 | 351779 |  |  |  |  |  |  |  |  |  |  |
| 5 | 589549 | 396862 |  |  |  |  |  |  |  |  |  |  |
| 6 | 532890 | 503601 |  |  |  |  |  |  |  |  |  |  |
| 7 | 340155 | 431014 |  |  |  |  |  |  |  |  |  |  |
| 8 | 269962 | 261959 |  |  |  |  |  |  |  |  |  |  |
| 9 | 170373 | 188950 |  |  |  |  |  |  |  |  |  |  |
| 10 | 94778 | 138143 |  |  |  |  |  |  |  |  |  |  |
| 11 | 33896 | 59211 |  |  |  |  |  |  |  |  |  |  |
| 12 | 24420 | 51090 |  |  |  |  |  |  |  |  |  |  |

Table 8.7.1.3. NE Atlantic Mackerel. WEIGHTS AT AGE IN THE CATCH

```
Units : Kg
    year
age 1980 1981 1982 1983 1984 1985 19 1986 1987 1988
    0.057 0.060 0.053 0.050 0.031 0.055 0.039}00.076 0.055 0.049 0.085 0.068 0.051 0.061
```



```
    0.249 0.248 0.249 0.219 0.184 0.262 0.245 0.223 0.259 0.237}00.233 0.253 0.239 0.240
    0.285 0.287 0.285 0.276 0.295 0.357 0.335 0.318 0.323 0.320}0.30.336 0.327 0.333 0.317
```



```
    0.378 0.377 0.378 0.386 0.344 0.417 0.471 0.474 0.456 0.433 0.423 0.423 0.460 0.436
    0.454 0.454 0.454 0.425 0.431 0.436 0.444 0.512 0.524 0.456 0.467 0.469 0.495 0.483
    0.498 0.499 0.496 0.435 0.542 0.521 0.457 0.493 0.555 0.543 0.528 0.506 0.532 0.527
    0.520 0.513 0.513 0.498 0.480 0.555 0.543 0.498 0.555 0.592 0.552 0.554 0.555 0.548
```




```
    1 0.590 0.576 0.574 0.608 0.636 0.679 0.694 0.635 0.624 0.648
    12 0.580 0.584 0.582 0.614 0.663 0.710 0.688 0.718 0.697 0.739 0.713 0.708 0.669 0.679
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007
    0}00.046 0.072 0.058 0.076 0.065 0.062 0.063 0.069 0.052 0.081 0.067 0.048 0.038 0.089
    1 0.136 0.143 0.143 0.143 0.157 0.176 0.135 0.172 0.160}00.170 0.156 0.151 0.071 0.120
    2 0.255 0.234 0.226 0.230 0.227 0.235 0.227 0.224 0.256 0.267 0.263 0.268 0.197 0.215
    0.339 0.333 0.313 0.295 0.310 0.306 0.306 0.305 0.307 0.336 0.323 0.306 0.307 0.292
```



```
    5
    6
    7 0.543 0.539 0.518 0.481 0.462 0.500 0.501 0.496 0.512 0.522 0.519 0.496 0.494 0.512
```






```
    12 0.713 0.672 0.670 0.636 0.631 0.687 0.644 0.666 0.665 0.715 0.684 0.655 0.689 0.666
        year
age 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017
    0}00.051 0.104 0.048 0.029 0.089 0.091 0.043 0.051 0.035 0.018
    1}00.105 0.153 0.118 0.113 0.123 0.173 0.127 0.154 0.158 0.178
    0.222 0.213 0.221 0.231 0.187 0.234 0.232 0.242 0.240 0.266
    0.292 0.283 0.291 0.282 0.285 0.277 0.282 0.294 0.297 0.312
```




```
    6}00.444 0.424 0.418 0.411 0.401 0.386 0.395 0.392 0.383 0.397
    0.497 0.450 0.471 0.451 0.431 0.406 0.422 0.420 0.411 0.415
    8
```



```
    10}00.620 0.586 0.573 0.580 0.537 0.472 0.482 0.489 0.479 0.484
    11 0.595 0.599 0.604 0.611 0.538 0.493 0.523 0.522 0.499 0.497
    12 0.662 0.630 0.630}0.6.664 0.585 0.554 0.583 0.560 0.520 0.531
```

Table 8.7.1.4. NE Atlantic Mackerel. WEIGHTS AT AGE IN THE STOCK

```
Units : Kg
    year
```



```
    0
    1
    0.205 0.179 0.159 0.179 0.204 0.244 0.184 0.157 0.181 0.162 0.166 0.184 0.201 0.190
    0.287 0.258 0.217 0.233 0.251 0.281 0.269 0.234 0.238 0.230}00.247 0.243 0.260 0.266
    0.322 0.312 0.300 0.282 0.293 0.308 0.301 0.318}00.298 0.272 0.290 0.303 0.308 0.323
    0.356 0.335 0.368 0.341 0.326 0.336 0.350}00.368 0.348 0.338 0.332 0.347 0.360 0.359
    6
    7 0.402 0.415 0.411 0.404 0.430 0.407 0.374 0.415 0.445 0.388 0.435 0.423 0.419 0.432
    8}00.434 0.431 0.456 0.438 0.455 0.455 0.434 0.431 0.442 0.449 0.447 0.492 0.458 0.459
    9}00.438 0.454 0.455 0.475 0.489 0.447 0.428 0.483 0.466 0.432 0.494 0.500 0.487 0.480
```




```
    12 0.534 0.531 0.544 0.528 0.567 0.591 0.542 0.581 0.594 0.556 0.536 0.615 0.568 0.577
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007
    0}0.0000.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1 0.114 0.116 0.097 0.084 0.083 0.087 0.093 0.113 0.109 0.112 0.112 0.106 0.108 0.083
    2 0.163 0.201 0.185 0.196 0.172 0.210 0.194 0.190 0.206 0.181 0.158 0.140 0.164 0.149
    3 0.240 0.278 0.250 0.257 0.248 0.260 0.253 0.246 0.245 0.251 0.258 0.221 0.236 0.206
    4 0.306 0.327 0.322 0.310}00.299 0.317 0.301 0.303 0.288 0.277 0.318 0.328 0.291 0.288
    5
    6}00.418 0.432 0.425 0.401 0.383 0.392 0.394 0.398 0.360 0.401 0.406 0.403 0.400 0.362
    7 0.459 0.458 0.446 0.460 0.409 0.424 0.416 0.417 0.418 0.407 0.449 0.464 0.413 0.451
    8}00.480 0.491 0.471 0.473 0.455 0.456 0.438 0.451 0.429 0.489 0.482 0.481 0.437 0.452
    9}00.496 0.511 0.513 0.505 0.475 0.489 0.464 0.484 0.458 0.490 0.506 0.547 0.455 0.508
    10}00.550 0.517 0.508 0.511 0.530 0.508 0.489 0.521 0.511 0.488 0.519 0.538 0.469 0.527
    11 0.592 0.560 0.538 0.546 0.500 0.545 0.514 0.535 0.523 0.521 0.579 0.509 0.531 0.533
    12 0.604 0.602 0.573 0.585 0.547 0.576 0.551 0.574 0.557 0.540 0.588 0.603 0.566 0.586
        year
age 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017
    0}0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1 0.133 0.107 0.096 0.080}0.0.089 0.076 0.107 0.078 0.059 0.058
    2 0.160 0.162 0.161 0.175 0.155 0.144 0.165 0.207 0.184 0.204
    3}00.207 0.214 0.201 0.223 0.216 0.179 0.199 0.247 0.239 0.237
    4 0.260}00.268 0.249 0.274 0.255 0.249 0.238 0.254 0.283 0.278
    5
    6}00.354 0.351 0.342 0.369 0.312 0.318 0.321 0.336 0.336 0.308
    70.393 0.386 0.389 0.389 0.360 0.341 0.341 0.350}00.364 0.338
    8
    9
```



```
    11 0.487 0.548 0.535 0.518 0.503 0.500 0.472 0.485 0.442 0.430
    12 0.511 0.559 0.573 0.525 0.557 0.520 0.517 0.549 0.470 0.494
```

Table 8.7.1.5. NE Atlantic Mackerel. NATURAL MORTALITY

```
Units : NA
    year
age 1980 1981 1982 1983 1984 1985 1986 1987 1988
```



```
    1
    2 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    3 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    4 0.15}0.1
    5
    6
    7 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```





```
    11
    12 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
        year
age 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007 2008 2009 2010 2011 2012 2013
    0}00.1
```



```
    2 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```



```
    4
    5
    6}00.1
    7 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```





```
    11 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    12 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
        year
age 2014 2015 2016 2017
    0}00.15\quad0.15\quad0.15\quad0.1
    1 0.15 0.15 0.15 0.15
    2 0.15 0.15 0.15 0.15
    3}00.150.15 0.15 0.1
    4
    5
    6}00.15\quad0.15 0.15 0.15
    70.15}00.1
    8
    9
    10}00.15\quad0.15 0.15 0.1
    11 0.15 0.15 0.15 0.15
    12 0.15 0.15 0.15 0.15
```

Table 8.7.1.6. NE Atlantic Mackerel. PROPORTION MATURE

```
    year
age 1980
    00.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1
    2
    3 0.872 0.837 0.857 0.863 0.853 0.885 0.926 0.922 0.901 0.915 0.909 0.912 0.928 0.934
    4 0.949 0.934 0.930 0.940 0.938 0.940 0.983 0.994 0.989 0.994 0.996 0.991 0.996 0.996
```



```
    6
    7 0.990 0.987 0.985 0.984 0.975 0.976 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    8 1.000 0.999 0.999 0.999 0.999 0.999 0.991 0.992 0.991 0.993 0.995 1.000 1.000 1.000
    9 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    10 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    11 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    12 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007 
    0 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1
```



```
    3 0.938 0.931 0.936 0.880 0.886 0.876 0.909 0.909 0.937 0.940 0.962 0.959 0.928 0.921
```



```
    5
    6
    7 0.999 0.999 0.999 1.000 1.000 1.000 1.000 0.999 1.000 0.999 0.999 0.999 1.000 1.000
    8 1.000 1.000 1.000 0.994 0.995 0.996 0.997 0.997 1.000 1.000 1.000 1.000 1.000 1.000
    9 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    10 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    11 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    12 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
        year
age 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017
    0 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1}00.095 0.096 0.096 0.096 0.094 0.092 0.092 0.104 0.103 0.101
    2 0.524 0.541 0.667 0.655 0.604 0.683 0.675 0.763 0.755 0.749
    3 0.917 0.919 0.930 0.927 0.926 0.921 0.916 0.944 0.941 0.936
    4 0.999}00.999 0.999 0.999 0.999 0.998 0.999 0.998 0.998 0.998
    5 0.999 1.000 1.000 1.000 0.999 1.000 1.000 0.999 0.999 1.000
    6 1.000 1.000 0.999 0.999 0.999 0.999 0.999 1.000 1.000 1.000
    7 1.000 1.000 1.000 1.000 1.000 1.000 0.999 0.999 0.999 0.999
    8 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    9 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    10 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    11 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    12 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
```

Table 8.7.1.7. NE Atlantic Mackerel. FRACTION OF HARVEST BEFORE SPAWNING

```
    year
age 1980
    0}0.0000.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1
    2
    3 0.222 0.222 0.222 0.222 0.222 0.222 0.222 0.222 0.222 0.222 0.253 0.285}00.316 0.318
    4 0.222 0.222 0.222 0.222 0.222 0.222 0.222 0.222 0.222 0.222 0.253 0.285 0.316 0.318
    5 0.381 0.381 0.381 0.381 0.381 0.381 0.381
    6
    7
    8}00.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.392 0.403 0.414 0.439
    9 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.392 0.403 0.414 0.439
    10}00.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.392 0.403 0.414 0.439
    11 0.381 0.381 0.381 0.381 0.381 
    l2 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.381 0.392 0.403 0.414 0.439
        year
    age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007
    0 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    1
    0.252 0.287 0.250 0.212 0.175 0.179 0.183 0.187 0.201 0.216 0.231 0.230}00.229 0.229
    0.321 0.323 0.328 0.334 0.339 0.364 0.390 0.415 0.408 0.400 0.393 0.375 0.357 0.338
    0.321 0.323 0.328 0.334 0.339 0.364 0.390 0.415 0.408 0.400 0.393 0.375 0.357 0.338
    0.464 0.489 0.492 0.494 0.497 0.462 0.425 0.390 0.405 0.420 0.434 0.402 0.368 0.336
```



```
    llllllllllllllllllllllll
    0.464 0.489 0.492 0.494 0.497 0.462 0.425 0.390 0.405 0.420 0.434 0.402 0.368 0.336
    0.464 0.489 0.492 0.494 0.497 0.462 0.425 0.390 0.405 0.420 0.434 0.402 0.368 0.336
```





```
        year
    age 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017
        0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
        llllllllllllll
        0.197 0.165 0.133 0.126 0.119}0.111 0.137 0.164 0.191 0.191
        0.305 0.270 0.237 0.183 0.129 0.075 0.121 0.168 0.214 0.214
        0.305 0.270 0.237 0.183 0.129 0.075 0.121 0.168 0.214 0.214
```



```
        0.305 0.272 0.241}00.232 0.223 0.214 0.199 0.183 0.169 0.169
        0.305 0.272 0.241 0.232 0.223 0.214 0.199 0.183 0.169 0.169
        0.305 0.272 0.241 0.232 0.223 0.214 0.199 0.183 0.169 0.169
        0.305 0.272 0.241 0.232 0.223}00.214 0.199 0.183 0.169 0.169
        0.305 0.272 0.241 0.232 0.223}00.214 0.199 0.183 0.169 0.169
```



```
    12 0.305 0.272 0.241 0.232 0.223 0.214 0.199 0.183 0.169 0.169
```

Table 8.7.1.8. NE Atlantic Mackerel. FRACTION OF NATURAL MORTALITY BEFORE SPAWNING

```
    year
age 1980 1981 1982 1983 1984 1985 1986 1987 1988 1989 1990
    0 0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345 0.333 0.341
    1}00.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345 0.333 0.341
    2}00.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345 0.333 0.341
```



```
    4 0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345 0.333 0.341
    5 0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345 0.333 0.341
    6
    7}00.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345 0.333 0.341
    8
    9 0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345 0.333 0.341
    10 0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345 0.333 0.341
    11 0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388}0.3.378 0.369 0.357 0.345 0.333 0.341
    12
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007
    0}00.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355 0.350 0.346 0.342 0.339
    1}00.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355 0.350 0.346 0.342 0.339
    2
    3}00.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355 0.350 0.346 0.342 0.339
    4 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355 0.350 0.346 0.342 0.339
    5 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355 0.350 0.346 0.342 0.339
    6
    7
    8}00.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355 0.350 0.346 0.342 0.339
    9 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355 0.350 0.346 0.342 0.339
    10 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355 0.350 0.346 0.342 0.339
    11 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355 0.350}00.346 0.342 0.339
    12
        year
    age 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017
    0}00.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311 0.343 0.343
    1
    2
    3}00.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311 0.343 0.343
    4 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311 0.343 0.343
    5
    6}00.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311 0.343 0.343
    7}00.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311 0.343 0.343
    8}00.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311 0.343 0.343
    9 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311 0.343 0.343
    10}00.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311 0.343 0.343
    11}00.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311 0.343 0.343
    12 0.311 0.283 0.255 0.252 0.249}00.246 0.278 0.311 0.343 0.343
```

Table 8.7.1.9. NE Atlantic Mackerel. SURVEY INDICES

```
Som
SSB-egg-based-survey
\begin{tabular}{ll}
1992 & 2018 \\
1 & 1
\end{tabular}
-1 ll
13874476.93
-1
-1
3766378.516
-1
-1
4198626.531
-1
-1
3233833.244
-1
-1
3106808.703
-1
-1
3782966.707
-1
-1
4810751.571
-1
-1
4831948.353
-1
-1
3524054.85
-1
-1
R-idx(sqrt transf)
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|}
\hline 1998 & 2017 & & & & & & & & \\
\hline 1 & 1 & 0 & 0 & & & & & & \\
\hline 0 & 0 & & & & & & & & \\
\hline 1 & 0.015720 & 899 & & & & & & & \\
\hline 1 & 0.017996 & 206 & & & & & & & \\
\hline 1 & 0.012743 & 674 & & & & & & & \\
\hline 1 & 0.022164 & 525 & & & & & & & \\
\hline 1 & 0.023618 & 634 & & & & & & & \\
\hline 1 & 0.013230 & 785 & & & & & & & \\
\hline 1 & 0.024607 & 411 & & & & & & & \\
\hline 1 & 0.038156 & 211 & & & & & & & \\
\hline 1 & 0.037598 & 707 & & & & & & & \\
\hline 1 & 0.020352 & 249 & & & & & & & \\
\hline 1 & 0.018292 & 615 & & & & & & & \\
\hline 1 & 0.015170 & 405 & & & & & & & \\
\hline 1 & 0.027764 & 032 & & & & & & & \\
\hline 1 & 0.036979 & 005 & & & & & & & \\
\hline 1 & 0.024205 & & & & & & & & \\
\hline 1 & 0.023257 & 095 & & & & & & & \\
\hline 1 & 0.025778 & 066 & & & & & & & \\
\hline 1 & 0.023169 & 671 & & & & & & & \\
\hline 1 & -1 & & & & & & & & \\
\hline 1 & -1 & & & & & & & & \\
\hline Swept & & & & & & & & & \\
\hline 2010 & 2018 & & & & & & & & \\
\hline 1 & 1 & 0.58 & 0.75 & & & & & & \\
\hline 3 & 11 & & & & & & & & \\
\hline 1 & 1617005 & 4035646 & 3059146 & 1591100 & 691936 & 413253 & 198106 & 65803 & 24747 \\
\hline 1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 \\
\hline 1 & 1283247 & 2383260 & 2164365 & 2850847 & 1783942 & 740361 & 299490 & 149282 & 84344 \\
\hline 1 & 9201746 & 2456618 & 3073772 & 3218990 & 2540444 & 1087937 & 377406 & 144695 & 146826 \\
\hline 1 & 7034162 & 4896456 & 2659443 & 2630617 & 2768227 & 1910160 & 849010 & 379745 & 95304 \\
\hline 1 & 2539963 & 6409324 & 4802298 & 1795564 & 1628872 & 1254859 & 727691 & 270562 & 72410 \\
\hline 1 & 1374705 & 2635033 & 5243607 & 4368491 & 1893026 & 1658839 & 1107866 & 754993 & 450100 \\
\hline 1 & 3562908 & 1953609 & 3318099 & 4680603 & 4653944 & 1754954 & 1944991 & 626406 & 507546 \\
\hline 1 & 496595 & 2384310 & 1200541 & 1408582 & 2330520 & 1787503 & 1049868 & 499295 & 557573 \\
\hline
\end{tabular}
```

Table 8.7.2.1. NE Atlantic Mackerel. SAM parameter estimates for the 2018 update.

|  | estimate | std. <br> dev | confidence interval lower bound | confidence interval upper bound |
| :---: | :---: | :---: | :---: | :---: |
| observation standard deviations |  |  |  |  |
| Catches | 0.18 | 0.15 | 0.14 | 0.25 |
| Egg survey | 0.20 | 0.29 | 0.11 | 0.37 |
| Recruitment index | 0.36 | 0.23 | 0.23 | 0.56 |
| IESSNS age 3 | 0.68 | 0.27 | 0.40 | 1.15 |
| IESSNS ages 4-11 | 0.40 | 0.20 | 0.27 | 0.59 |
| Recapture overdispersion steal tags | 1.21 | 0.27 | 1.36 | 1.12 |
| Recapture overdispersion RFID tags | 1.16 | 0.63 | 1.55 | 1.04 |
| random walk standard deviation |  |  |  |  |
| F | 0.25 | 0.15 | 0.18 | 0.33 |
| N@age0 | 0.78 | 0.15 | 0.58 | 1.06 |
| process error standard deviation |  |  |  |  |
| N@age1-12+ | 0.17 | 0.13 | 0.13 | 0.21 |
| catchabilities |  |  |  |  |
| egg survey | 1.37 | 0.08 | 1.16 | 1.61 |
| recruitment index | 0.00 | 0.12 | 0.00 | 0.00 |
| IESSNS age 3 | 1.00 | 0.27 | 0.58 | 1.71 |
| IESSNS age 4 | 1.49 | 0.18 | 1.04 | 2.14 |
| IESSNS age 5 | 1.99 | 0.18 | 1.39 | 2.85 |
| IESSNS age 6 | 2.35 | 0.18 | 1.63 | 3.38 |
| IESSNS age 7 | 2.69 | 0.18 | 1.87 | 3.87 |
| IESSNS age 8 | 2.57 | 0.18 | 1.77 | 3.71 |
| IESSNS age 9 | 2.56 | 0.18 | 1.77 | 3.70 |
| IESSNS ages 10-11 | 2.18 | 0.18 | 1.52 | 3.12 |
| post tagging survival steal tags | 0.39 | 0.10 | 0.34 | 0.43 |
| post tagging survival RFID tags | 0.10 | 0.08 | 0.09 | 0.12 |

Table 8.7.3.1. NE Atlantic Mackerel. STOCK SUMMARY. Low = lower limit and High = higher limit of 95\% confidence interval.

| Year | Recruitment (Age 0) | High | Low | SSB | High | Low | Total Catch | F (Ages 4-8) | High | Low |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | THOUSANDS |  |  | TONNES |  |  | tonnes | PER YEAR |  |  |
| 1980 | 7750521 | 16984113 | 3536869 | 4017907 | 8457837 | 1908712 | 734950 | 0.171 | 0.34 | 0.087 |
| 1981 | 6406269 | 12337825 | 3326379 | 3639690 | 6906887 | 1917990 | 754045 | 0.171 | 0.32 | 0.091 |
| 1982 | 1976069 | 4143932 | 942305 | 3651751 | 6260529 | 2130057 | 716987 | 0.172 | 0.31 | 0.095 |
| 1983 | 1571022 | 3517044 | 701756 | 3969695 | 6114088 | 2577404 | 672283 | 0.173 | 0.30 | 0.099 |
| 1984 | 5911986 | 11635802 | 3003796 | 4194238 | 6089443 | 2888873 | 641928 | 0.174 | 0.29 | 0.104 |
| 1985 | 3856995 | 7251430 | 2051514 | 4034958 | 5635108 | 2889188 | 614371 | 0.179 | 0.29 | 0.110 |
| 1986 | 3835380 | 6929670 | 2122776 | 3661554 | 4952706 | 2707000 | 602201 | 0.186 | 0.29 | 0.118 |
| 1987 | 5394520 | 9378868 | 3102810 | 3689032 | 4948576 | 2750075 | 654992 | 0.195 | 0.30 | 0.127 |
| 1988 | 3362579 | 5796405 | 1950681 | 3609380 | 4717758 | 2761401 | 680491 | 0.20 | 0.30 | 0.138 |
| 1989 | 3601083 | 6225603 | 2082979 | 3334286 | 4254397 | 2613169 | 585920 | 0.22 | 0.32 | 0.153 |
| 1990 | 2584288 | 4615319 | 1447038 | 3390278 | 4220794 | 2723181 | 626107 | 0.25 | 0.35 | 0.174 |
| 1991 | 3243249 | 5537959 | 1899375 | 3226020 | 3954919 | 2631459 | 675665 | 0.28 | 0.39 | 0.20 |
| 1992 | 3886031 | 6639093 | 2274594 | 2890322 | 3483688 | 2398022 | 760690 | 0.31 | 0.43 | 0.23 |
| 1993 | 3045153 | 5191337 | 1786237 | 2526363 | 3023607 | 2110892 | 824568 | 0.35 | 0.47 | 0.26 |
| 1994 | 2888409 | 4882393 | 1708774 | 2202704 | 2619309 | 1852360 | 819087 | 0.36 | 0.49 | 0.27 |
| 1995 | 2425238 | 4122421 | 1426778 | 2198832 | 2596179 | 1862300 | 756277 | 0.34 | 0.44 | 0.26 |
| 1996 | 3468432 | 6357582 | 1892232 | 2092064 | 2458667 | 1780123 | 563472 | 0.29 | 0.38 | 0.23 |
| 1997 | 2676317 | 4716590 | 1518612 | 2078029 | 2412438 | 1789976 | 573029 | 0.27 | 0.35 | 0.21 |
| 1998 | 3246924 | 5010572 | 2104054 | 2109062 | 2458101 | 1809586 | 666316 | 0.27 | 0.35 | 0.21 |
| 1999 | 3753406 | 5630541 | 2502079 | 2253952 | 2610229 | 1946305 | 640309 | 0.30 | 0.37 | 0.24 |
| 2000 | 2588498 | 3794164 | 1765955 | 2181219 | 2482469 | 1916526 | 738606 | 0.33 | 0.39 | 0.29 |
| 2001 | 5132275 | 7328207 | 3594365 | 2059605 | 2327376 | 1822643 | 737463 | 0.39 | 0.45 | 0.33 |
| 2002 | 8708579 | 12577632 | 6029700 | 1885840 | 2146182 | 1657079 | 771422 | 0.43 | 0.50 | 0.37 |
| 2003 | 2992952 | 4284375 | 2090798 | 1910203 | 2198387 | 1659796 | 679287 | 0.46 | 0.54 | 0.39 |
| 2004 | 3936957 | 5726238 | 2706774 | 2410549 | 2831564 | 2052132 | 660491 | 0.42 | 0.49 | 0.36 |


| Year | Recruitment (Age 0) | High | Low | SSB | High | Low | Total Catch | F (Ages 4-8) | High | Low |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | THOUSANDS |  |  | TONNES |  |  | tonnes | PER YEAR |  |  |
| 2005 | 6300180 | 9244002 | 4293840 | 2269837 | 2705129 | 1904590 | 549514 | 0.32 | 0.38 | 0.27 |
| 2006 | 11464574 | 16692197 | 7874126 | 2215900 | 2618927 | 1874894 | 481181 | 0.29 | 0.34 | 0.25 |
| 2007 | 5374061 | 7823457 | 3691531 | 2417312 | 2820632 | 2071663 | 586206 | 0.33 | 0.39 | 0.28 |
| 2008 | 5604311 | 8168336 | 3845128 | 2986019 | 3534299 | 2522795 | 623165 | 0.32 | 0.38 | 0.27 |
| 2009 | 5200071 | 7550819 | 3581167 | 3634054 | 4324811 | 3053624 | 737969 | 0.29 | 0.34 | 0.25 |
| 2010 | 6683926 | 9631467 | 4638427 | 4025533 | 4732936 | 3423862 | 875515 | 0.28 | 0.34 | 0.24 |
| 2011 | 7483547 | 10773616 | 5198207 | 4794839 | 5605302 | 4101560 | 946661 | 0.28 | 0.33 | 0.24 |
| 2012 | 4793523 | 6980496 | 3291723 | 4388467 | 5125670 | 3757293 | 892353 | 0.26 | 0.32 | 0.22 |
| 2013 | 3220460 | 4867764 | 2130621 | 4097288 | 4816614 | 3485388 | 931732 | 0.29 | 0.35 | 0.24 |
| 2014 | 8120609 | 12377522 | 5327746 | 4130139 | 4869649 | 3502932 | 1393000 | 0.33 | 0.40 | 0.27 |
| 2015 | 2588980 | 4292931 | 1561361 | 3962603 | 4726385 | 3322248 | 1208990 | 0.34 | 0.42 | 0.28 |
| 2016 | 784490 | 1514562 | 406338 | 3527235 | 4358303 | 2854640 | 1094066 | 0.34 | 0.43 | 0.26 |
| 2017 | 5267776* |  |  | 3081442 | 4048464 | 2345404 | 1155944 | 0.38 | 0.52 | 0.28 |
| 2018 | 3977184** |  |  | 2353927*** |  |  |  |  |  |  |

* Time-tapered weighted mean of recruitment estimates for 1990-2016.
** Geometric mean 1990-2016.
*** Estimated value from the forecast.

Table 8.7.3.2. NE Atlantic Mackerel. ESTIMATED POPULATION ABUNDANCE

| year |  | ands |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| 0 | 7750521 | 6406269 | 1976069 | 1571022 | 5911986 | 3856995 | 3835380 | 5394520 | 3362579 |
| 1 | 4653503 | 6674218 | 5704000 | 1616192 | 1251155 | 5420373 | 3217029 | 3214005 | 4790054 |
| 2 | 1994678 | 3851943 | 5596469 | 4948040 | 1288926 | 972292 | 4842709 | 2622680 | 2635582 |
| 3 | 859737 | 1618606 | 3202496 | 4822696 | 4516385 | 1006458 | 755269 | 4369121 | 2102982 |
| 4 | 1436702 | 677785 | 1255507 | 2669203 | 4073668 | 3961881 | 807785 | 568199 | 3770464 |
| 5 | 3133512 | 1086645 | 502708 | 900550 | 2047940 | 3175475 | 3039200 | 652426 | 394514 |
| 6 | 2510612 | 2319913 | 814036 | 382695 | 652194 | 1561824 | 2343384 | 2157767 | 523863 |
| 7 | 863764 | 1795418 | 1654449 | 581517 | 279837 | 470366 | 1093914 | 1620849 | 1453406 |
| 8 | 331686 | 616978 | 1282244 | 1179309 | 411693 | 205305 | 328654 | 786281 | 1123879 |
| 9 | 893401 | 236932 | 440339 | 915906 | 839983 | 294981 | 148484 | 228410 | 559979 |
| 10 | 254857 | 638392 | 169181 | 314231 | 654130 | 597984 | 212258 | 105041 | 157144 |
| 11 | 370675 | 182078 | 455649 | 120858 | 224335 | 466509 | 423241 | 149601 | 73519 |
| 12 | 720141 | 779558 | 686085 | 813009 | 664842 | 633036 | 776749 | 837345 | 681914 |
| year |  |  |  |  |  |  |  |  |  |
| age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| 0 | 3601083 | 2584288 | 3243249 | 3886031 | 3045153 | 2888409 | 2425238 | 3468432 | 2676317 |
| 1 | 2791294 | 3116740 | 2146980 | 2776509 | 3371213 | 2574073 | 2473456 | 2018977 | 3033938 |
| 2 | 4151447 | 2263634 | 2630428 | 1741654 | 2320685 | 2853235 | 2120474 | 2063911 | 1638484 |
| 3 | 2267428 | 3809144 | 1986829 | 2416739 | 1482507 | 1901314 | 2351623 | 2087839 | 1886939 |
| 4 | 1648739 | 1800686 | 2923925 | 1458623 | 1936356 | 1048497 | 1395745 | 1761690 | 1728231 |
| 5 | 2885724 | 1094092 | 1239644 | 1838572 | 946709 | 1315821 | 667912 | 947183 | 1183072 |
| 6 | 280542 | 1993896 | 803823 | 946094 | 1156693 | 598254 | 934839 | 486344 | 700631 |
| 7 | 411136 | 188509 | 1271565 | 511140 | 595837 | 667257 | 364148 | 563023 | 323875 |
| 8 | 1060679 | 309038 | 123601 | 745711 | 311527 | 325256 | 282228 | 208508 | 331667 |
| 9 | 779546 | 719709 | 216105 | 77710 | 403224 | 169709 | 159252 | 135522 | 141769 |
| 10 | 388105 | 513918 | 474639 | 135206 | 43802 | 198426 | 90147 | 81118 | 84077 |
| 11 | 105748 | 262315 | 324277 | 284713 | 77335 | 22471 | 103818 | 48353 | 44956 |
| 12 | 516017 | 416934 | 438303 | 459432 | 412102 | 261141 | 145339 | 134326 | 108136 |
| year |  |  |  |  |  |  |  |  |  |
| age | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 0 | 3246924 | 3753406 | 2588498 | 5132275 | 8708579 | 2992952 | 3936957 | 6300180 | 11464574 |
| 1 | 2230568 | 2717670 | 3214851 | 1656113 | 5540645 | 7640107 | 2374435 | 3434107 | 6745432 |
| 2 | 2589748 | 1811668 | 2225301 | 2598715 | 1156202 | 4778023 | 6322183 | 2163769 | 3206972 |
| 3 | 1235950 | 2399436 | 1606592 | 1718343 | 2288967 | 863689 | 3630261 | 4805540 | 1796105 |
| 4 | 1603085 | 1148736 | 1825073 | 1181814 | 1376087 | 1507762 | 725866 | 2007947 | 3141256 |
| 5 | 1442468 | 1246929 | 887036 | 1248500 | 872918 | 828969 | 939853 | 533921 | 1174809 |
| 6 | 843866 | 899401 | 858755 | 571109 | 792513 | 513453 | 441021 | 494521 | 374910 |
| 7 | 466220 | 596413 | 617614 | 568121 | 353005 | 384012 | 253679 | 230731 | 283550 |
| 8 | 253753 | 305369 | 371428 | 407934 | 323896 | 207701 | 173905 | 124998 | 134023 |
| 9 | 207980 | 177515 | 189592 | 233962 | 215643 | 170164 | 103054 | 84052 | 70411 |
| 10 | 97984 | 132390 | 114796 | 116459 | 121321 | 104783 | 77607 | 54332 | 49742 |
| 11 | 54502 | 64192 | 75161 | 69488 | 59571 | 58960 | 43288 | 31887 | 30210 |
| 12 | 101486 | 106159 | 114888 | 117007 | 101807 | 74253 | 52784 | 40653 | 38505 |
| year |  |  |  |  |  |  |  |  |  |
| age | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| 0 | 5374061 | 5604311 | 5200071 | 6683926 | 7483547 | 4793523 | 3220460 | 8120609 | 2588980 |
| 1 | 8364901 | 5023296 | 4767259 | 5132389 | 6100878 | 6435719 | 3666510 | 2653561 | 5729227 |
| 2 | 5761560 | 6799560 | 4475215 | 4861233 | 4241194 | 5854904 | 5875273 | 2836051 | 2139734 |
| 3 | 2674246 | 4974795 | 5957253 | 4072078 | 4872581 | 3228525 | 5167586 | 5056793 | 2190824 |
| 4 | 1524897 | 2160727 | 4082726 | 4939563 | 3665559 | 3604041 | 2320908 | 3809943 | 3546703 |
| 5 | 2021532 | 1211576 | 1694921 | 2990105 | 3640896 | 2898278 | 2438039 | 1820046 | 2697023 |
| 6 | 756155 | 1134335 | 857849 | 1261783 | 2099677 | 2599148 | 2061105 | 1749681 | 1337925 |
| 7 | 252284 | 432420 | 656180 | 546902 | 858589 | 1388862 | 1516822 | 1486047 | 1268619 |
| 8 | 171286 | 161852 | 249913 | 348960 | 366556 | 552617 | 755650 | 1005593 | 1017669 |
| 9 | 87865 | 91768 | 100362 | 153527 | 193504 | 232924 | 321638 | 472405 | 684094 |
| 10 | 42040 | 51283 | 47692 | 65513 | 85207 | 115615 | 126056 | 204054 | 330069 |
| 11 | 30128 | 20710 | 27323 | 24875 | 40456 | 48998 | 70568 | 69010 | 105121 |
| 12 | 37494 | 30676 | 22847 | 30202 | 34270 | 41879 | 50179 | 53444 | 68892 |
| year |  |  |  |  |  |  |  |  |  |
| age | 2016 | 2017 | 2018 |  |  |  |  |  |  |
| 0 | 784490 | 12549868 | 12549868 |  |  |  |  |  |  |
| 1 | 1945108 | 832500 | 10779246 |  |  |  |  |  |  |
| 2 | 4183158 | 1544466 | 685512 |  |  |  |  |  |  |
| 3 | 1581349 | 3327781 | 1156774 |  |  |  |  |  |  |
| 4 | 1786534 | 1275182 | 2501606 |  |  |  |  |  |  |
| 5 | 2338797 | 1337856 | 831220 |  |  |  |  |  |  |
| 6 | 1814684 | 1584788 | 788762 |  |  |  |  |  |  |
| 7 | 909373 | 1282719 | 965217 |  |  |  |  |  |  |
| 8 | 819252 | 622176 | 716721 |  |  |  |  |  |  |
| 9 | 557457 | 560140 | 389037 |  |  |  |  |  |  |
| 10 | 358043 | 339104 | 258445 |  |  |  |  |  |  |
| 11 | 174934 | 214525 | 221993 |  |  |  |  |  |  |
| 12 | 86911 | 155190 | 205492 |  |  |  |  |  |  |

Table 8.7.3.3. NE Atlantic Mackerel. ESTIMATED FISHING MORTALITY

```
    year
age 1980
    0}0.0100.010 0.010 0.010 0.010 0.011 0.010 0.010 0.011 0.011 0.011 0.011 0.011 0.011
    1
    2
    3 0.086 0.086 0.085 0.085 0.086 0.088 0.091 0.094 0.098 0.102 0.107 0.112 0.116 0.121
    4 0.148 0.148 0.149 0.149 0.150 0.154 0.160 0.173 0.184 0.201 0.215 0.229 0.234 0.234
    5
    6
    7}00.186 0.186 0.186 0.185 0.186 0.190 0.198 0.209 0.221 0.240 0.278 0.339 0.409 0.479
    8 0.186 0.186 0.186 0.185 0.186 0.190 0.198 0.209 0.221 0.240 0.278 0.339 0.409 0.479
    9 0.186 0.186 0.186 0.185 0.186 0.190 0.198 0.209 0.221 0.240 0.278 0.339 0.409 0.479
```





```
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007
    0}00.011 0.011 0.011 0.011 0.011 0.012 0.012 0.007 0.007 0.005 0.004 0.003 0.005 0.004
```



```
    0.054 0.055 0.057 0.060 0.062 0.065 0.069 0.068 0.068 0.068 0.073 0.066 0.055 0.039
    0.124 0.124 0.125 0.128 0.136 0.154 0.177 0.165 0.169 0.142 0.159 0.146 0.109 0.096
    0.233 0.224 0.215 0.203 0.207 0.223 0.254 0.281 0.289}0.2.251 0.229 0.193 0.180 0.166
    0.222 0.223 0.231 0.249 0.278 0.306 0.357 0.338}00.367 0.368 0.343 0.279 0.240 0.265
    0.318}00.311 0.296 0.290 0.291 0.313 0.357 0.438 0.443 0.464 0.425 0.350 0.329 0.332
```



```
    0.520 0.463 0.362 0.294 0.290 0.320 0.353 0.437 0.523 0.611 0.549 0.387 0.353 0.456
```






```
        year
    age 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017
        0.005 0.004 0.004 0.004 0.005 0.004 0.005 0.003 0.002 0.002
        0.012 0.010 0.013 0.012 0.012 0.014 0.017 0.021 0.028 0.044
        0.030}00.029 0.031 0.032 0.037 0.044 0.052 0.059 0.057 0.061
        0.090 0.086 0.082 0.078 0.081 0.095 0.131 0.139 0.166 0.164
        0.160 0.169 0.167 0.154 0.154 0.189 0.227 0.219 0.267 0.308
        0.252 0.233 0.235 0.218}00.211 0.237 0.304 0.297 0.309 0.343
        0.307 0.307 0.282 0.278}00.263 0.275 0.345 0.379 0.361 0.389
        0.443 0.370 0.366 0.374 0.347 0.382 0.384 0.415 0.369 0.437
        0.443 0.370 0.366 0.374 0.347 0.382 0.384 0.415 0.369 0.437
        0.443 0.370}00.366 0.374 0.347 0.382 0.384 0.415 0.369 0.437
        0.443 0.370}00.366 0.374 0.347 0.382 0.384 0.415 0.369 0.437
        11}00.443 0.370 0.366 0.374 0.347 0.382 0.384 0.415 0.369 0.437
        12 0.443 0.370 0.366 0.374 0.347 0.382 0.384 0.415 0.369 0.437
```

Table 8.8.3.1. NE Atlantic Mackerel. Short-term prediction: INPUT DATA

|  |  | $\Sigma$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2018 |  |  |  |  |  |  |  |  |
| 0 | 3977184 | 0.15 | 0 | 0 | 0.332 | 0 | 0.002 | 0.035 |
| 1 | 4524562 | 0.15 | 0.10 | 0.182 | 0.332 | 0.065 | 0.031 | 0.163 |
| 2 | 685511.6 | 0.15 | 0.76 | 0.182 | 0.332 | 0.198 | 0.059 | 0.249 |
| 3 | 1156774 | 0.15 | 0.94 | 0.199 | 0.332 | 0.241 | 0.157 | 0.301 |
| 4 | 2501606 | 0.15 | 1 | 0.199 | 0.332 | 0.272 | 0.264 | 0.335 |
| 5 | 831220 | 0.15 | 1 | 0.174 | 0.332 | 0.298 | 0.316 | 0.361 |
| 6 | 788762 | 0.15 | 1 | 0.174 | 0.332 | 0.327 | 0.377 | 0.391 |
| 7 | 965217 | 0.15 | 1 | 0.174 | 0.332 | 0.351 | 0.407 | 0.415 |
| 8 | 716721 | 0.15 | 1 | 0.174 | 0.332 | 0.380 | 0.407 | 0.442 |
| 9 | 389037 | 0.15 | 1 | 0.174 | 0.332 | 0.403 | 0.407 | 0.461 |
| 10 | 258445 | 0.15 | 1 | 0.174 | 0.332 | 0.433 | 0.407 | 0.484 |
| 11 | 221993 | 0.15 | 1 | 0.174 | 0.332 | 0.452 | 0.407 | 0.506 |
| 12+ | 205492 | 0.15 | 1 | 0.174 | 0.332 | 0.504 | 0.407 | 0.537 |
| $2019$ |  |  |  |  |  |  |  |  |
| 0 | 3977184 | 0.15 | 0 | 0 | 0.332 | 0 | 0.002 | 0.035 |
| 1 | - | 0.15 | 0.10 | 0.182 | 0.332 | 0.065 | 0.031 | 0.163 |
| 2 | - | 0.15 | 0.76 | 0.182 | 0.332 | 0.198 | 0.059 | 0.249 |
| 3 | - | 0.15 | 0.94 | 0.199 | 0.332 | 0.241 | 0.157 | 0.301 |
| 4 | - | 0.15 | 1 | 0.199 | 0.332 | 0.272 | 0.264 | 0.335 |
| 5 | - | 0.15 | 1 | 0.174 | 0.332 | 0.298 | 0.316 | 0.361 |
| 6 | - | 0.15 | 1 | 0.174 | 0.332 | 0.327 | 0.377 | 0.391 |
| 7 | - | 0.15 | 1 | 0.174 | 0.332 | 0.351 | 0.407 | 0.415 |
| 8 | - | 0.15 | 1 | 0.174 | 0.332 | 0.380 | 0.407 | 0.442 |
| 9 | - | 0.15 | 1 | 0.174 | 0.332 | 0.403 | 0.407 | 0.461 |
| 10 | - | 0.15 | 1 | 0.174 | 0.332 | 0.433 | 0.407 | 0.484 |
| 11 | - | 0.15 | 1 | 0.174 | 0.332 | 0.452 | 0.407 | 0.506 |
| 12+ | - | 0.15 | 1 | 0.174 | 0.332 | 0.504 | 0.407 | 0.537 |
| 2020 |  |  |  |  |  |  |  |  |
| 0 | 3977184 | 0.15 | 0 | 0 | 0.332 | 0 | 0.002 | 0.035 |
| 1 | - | 0.15 | 0.10 | 0.182 | 0.332 | 0.065 | 0.031 | 0.163 |
| 2 | - | 0.15 | 0.76 | 0.182 | 0.332 | 0.198 | 0.059 | 0.249 |
| 3 | - | 0.15 | 0.94 | 0.199 | 0.332 | 0.241 | 0.157 | 0.301 |
| 4 | - | $0.15$ | 1 | $0.199$ | 0.332 | 0.272 | 0.264 | 0.335 |
| 5 | - | 0.15 | 1 | 0.174 | 0.332 | 0.298 | 0.316 | 0.361 |
| 6 | - | 0.15 | 1 | 0.174 | 0.332 | 0.327 | 0.377 | 0.391 |
| 7 | - | 0.15 | 1 | 0.174 | 0.332 | 0.351 | 0.407 | 0.415 |
| 8 | - | 0.15 | 1 | 0.174 | 0.332 | 0.380 | 0.407 | 0.442 |
| 9 | - | 0.15 | 1 | 0.174 | 0.332 | 0.403 | 0.407 | 0.461 |


|  |  | $\Sigma$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 | - | 0.15 | 1 | 0.174 | 0.332 | 0.433 | 0.407 | 0.484 |
| 11 | - | 0.15 | 1 | 0.174 | 0.332 | 0.452 | 0.407 | 0.506 |
| 12+ | - | 0.15 | 1 | 0.174 | 0.332 | 0.504 | 0.407 | 0.537 |

Table 8.8.3.2. NE Atlantic Mackerel. Short-term prediction: Multi-option table for $1001 \mathbf{k t}$ catch in 2018 and a range of F-values in 2019.

| $\mathbf{2 0 1 8}$ |  |  |  |
| :--- | :--- | :--- | :--- |
| TSB | SSB | Fbar | Catch |
| 2977734 | 2353927 | 0,455 | 1000559 |


| 2019 |  |  |  | 2020 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TSB | SSB | Fbar | Catch | TSB | SSB | Implied change in the catch |
|  |  |  |  |  |  |  |
| 2667210 | 2167164 | 0,00 | 0 | 3166783 | 2642785 | -100\% |
| - | 2164240 | 0,01 | 19776 | 3149804 | 2623332 | -98\% |
| - | 2161320 | 0,02 | 39374 | 3132980 | 2604087 | -96\% |
| - | 2158405 | 0,03 | 58798 | 3116309 | 2585049 | -94\% |
| - | 2155496 | 0,04 | 78048 | 3099789 | 2566214 | -92\% |
| - | 2152591 | 0,05 | 97126 | 3083420 | 2547581 | -90\% |
| - | 2149691 | 0,06 | 116034 | 3067199 | 2529148 | -88\% |
| - | 2146796 | 0,07 | 134774 | 3051124 | 2510910 | -87\% |
| - | 2143905 | 0,08 | 153347 | 3035196 | 2492867 | -85\% |
| - | 2141020 | 0,09 | 171755 | 3019411 | 2475016 | -83\% |
| - | 2138140 | 0,10 | 190001 | 3003769 | 2457355 | -81\% |
| - | 2135264 | 0,11 | 208084 | 2988267 | 2439882 | -79\% |
| - | 2132393 | 0,12 | 226007 | 2972906 | 2422593 | -77\% |
| - | 2129527 | 0,13 | 243772 | 2957682 | 2405488 | -76\% |
| - | 2126666 | 0,14 | 261380 | 2942596 | 2388563 | -74\% |
| - | 2123810 | 0,15 | 278833 | 2927644 | 2371817 | -72\% |
| - | 2120958 | 0,16 | 296132 | 2912827 | 2355248 | -70\% |
| - | 2118112 | 0,17 | 313279 | 2898142 | 2338853 | -69\% |
| - | 2115270 | 0,18 | 330275 | 2883589 | 2322630 | -67\% |
| - | 2112433 | 0,19 | 347122 | 2869166 | 2306578 | -65\% |
| - | 2109600 | 0,20 | 363822 | 2854871 | 2290695 | -64\% |
| - | 2106773 | 0,21 | 380375 | 2840704 | 2274977 | -62\% |
| - | 2103950 | 0,22 | 396784 | 2826663 | 2259424 | -60\% |
| - | 2101132 | 0,23 | 413049 | 2812746 | 2244034 | -59\% |
| - | 2098319 | 0,24 | 429172 | 2798953 | 2228804 | -57\% |
| - | 2095510 | 0,25 | 445155 | 2785283 | 2213732 | -56\% |
| - | 2092706 | 0,26 | 460999 | 2771733 | 2198818 | -54\% |
| - | 2089907 | 0,27 | 476705 | 2758303 | 2184058 | -52\% |
| - | 2087113 | 0,28 | 492275 | 2744992 | 2169451 | -51\% |
| - | 2084323 | 0,29 | 507710 | 2731799 | 2154996 | -49\% |
| - | 2081538 | 0,30 | 523012 | 2718721 | 2140690 | -48\% |
| - | 2078758 | 0,31 | 538181 | 2705759 | 2126532 | -46\% |
| - | 2075982 | 0,32 | 553219 | 2692910 | 2112520 | -45\% |
| - | 2073211 | 0,33 | 568128 | 2680174 | 2098652 | -43\% |
| - | 2070445 | 0,34 | 582908 | 2667550 | 2084926 | -42\% |
| - | 2067683 | 0,35 | 597561 | 2655037 | 2071342 | -40\% |
| - | 2064926 | 0,36 | 612089 | 2642632 | 2057896 | -39\% |
| - | 2062174 | 0,37 | 626491 | 2630336 | 2044589 | -37\% |
| - | 2059426 | 0,38 | 640771 | 2618147 | 2031417 | -36\% |
| - | 2056683 | 0,39 | 654928 | 2606065 | 2018380 | -35\% |
| - | 2053944 | 0,40 | 668965 | 2594087 | 2005475 | -33\% |
| - | 2051210 | 0,41 | 682881 | 2582213 | 1992702 | -32\% |
| - | 2048481 | 0,42 | 696680 | 2570442 | 1980058 | -30\% |
| - | 2045756 | 0,43 | 710360 | 2558774 | 1967543 | -29\% |
| - | 2043036 | 0,44 | 723925 | 2547206 | 1955154 | -28\% |
| - | 2040320 | 0,45 | 737375 | 2535738 | 1942890 | -26\% |
| - | 2037609 | 0,46 | 750710 | 2524368 | 1930751 | -25\% |
| - | 2034903 | 0,47 | 763933 | 2513097 | 1918733 | -24\% |


| - | 2032201 | 0,48 | 777044 | 2501922 | 1906836 | -22\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| - | 2029503 | 0,49 | 790045 | 2490844 | 1895059 | -21\% |
| - | 2026810 | 0,50 | 802936 | 2479860 | 1883400 | -20\% |
| - | 2024122 | 0,51 | 815719 | 2468971 | 1871858 | -18\% |
| - | 2021438 | 0,52 | 828394 | 2458174 | 1860431 | -17\% |
| - | 2018758 | 0,53 | 840964 | 2447470 | 1849117 | -16\% |
| - | 2016084 | 0,54 | 853427 | 2436857 | 1837917 | -15\% |
| - | 2013413 | 0,55 | 865787 | 2426334 | 1826828 | -13\% |
| - | 2010747 | 0,56 | 878044 | 2415901 | 1815849 | -12\% |
| - | 2008086 | 0,57 | 890198 | 2405556 | 1804978 | -11\% |
| - | 2005429 | 0,58 | 902251 | 2395299 | 1794216 | -10\% |
| - | 2002776 | 0,59 | 914204 | 2385128 | 1783559 | -9\% |
| - | 2000128 | 0,60 | 926058 | 2375044 | 1773008 | -7\% |
| - | 1997484 | 0,61 | 937814 | 2365044 | 1762560 | -6\% |
| - | 1994845 | 0,62 | 949472 | 2355129 | 1752215 | -5\% |
| - | 1992210 | 0,63 | 961034 | 2345297 | 1741971 | -4\% |
| - | 1989580 | 0,64 | 972501 | 2335548 | 1731828 | -3\% |
| - | 1986954 | 0,65 | 983873 | 2325880 | 1721783 | -2\% |
| - | 1984332 | 0,66 | 995152 | 2316293 | 1711837 | -1\% |
| - | 1981715 | 0,67 | 1006338 | 2306786 | 1701987 | 1\% |
| - | 1979102 | 0,68 | 1017433 | 2297359 | 1692234 | 2\% |
| - | 1976494 | 0,69 | 1028436 | 2288010 | 1682574 | 3\% |
| - | 1973890 | 0,70 | 1039350 | 2278739 | 1673009 | 4\% |
| - | 1971290 | 0,71 | 1050175 | 2269545 | 1663535 | 5\% |
| - | 1968695 | 0,72 | 1060911 | 2260427 | 1654153 | 6\% |
| - | 1966104 | 0,73 | 1071561 | 2251384 | 1644861 | 7\% |
| - | 1963517 | 0,74 | 1082124 | 2242417 | 1635659 | 8\% |
| - | 1960934 | 0,75 | 1092601 | 2233523 | 1626545 | 9\% |
| - | 1958356 | 0,76 | 1102993 | 2224702 | 1617518 | 10\% |
| - | 1955783 | 0,77 | 1113302 | 2215954 | 1608577 | 11\% |
| - | 1953213 | 0,78 | 1123527 | 2207278 | 1599722 | 12\% |
| - | 1950648 | 0,79 | 1133670 | 2198673 | 1590951 | 13\% |
| - | 1948087 | 0,80 | 1143732 | 2190138 | 1582263 | 14\% |
| - | 1945531 | 0,81 | 1153712 | 2181673 | 1573657 | 15\% |
| - | 1942978 | 0,82 | 1163613 | 2173276 | 1565133 | 16\% |
| - | 1940430 | 0,83 | 1173435 | 2164949 | 1556689 | 17\% |
| - | 1937886 | 0,84 | 1183178 | 2156688 | 1548325 | 18\% |
| - | 1935347 | 0,85 | 1192843 | 2148495 | 1540040 | 19\% |
| - | 1932812 | 0,86 | 1202432 | 2140368 | 1531832 | 20\% |
| - | 1930281 | 0,87 | 1211944 | 2132307 | 1523701 | 21\% |
| - | 1927754 | 0,88 | 1221381 | 2124311 | 1515646 | 22\% |
| - | 1925231 | 0,89 | 1230744 | 2116379 | 1507667 | 23\% |
| - | 1922713 | 0,90 | 1240032 | 2108511 | 1499761 | 24\% |
| - | 1920199 | 0,91 | 1249247 | 2100706 | 1491929 | 25\% |
| - | 1917689 | 0,92 | 1258389 | 2092964 | 1484170 | 26\% |
| - | 1915183 | 0,93 | 1267460 | 2085283 | 1476482 | 27\% |
| - | 1912681 | 0,94 | 1276459 | 2077664 | 1468866 | 28\% |
| - | 1910184 | 0,95 | 1285388 | 2070106 | 1461319 | 28\% |
| - | 1907691 | 0,96 | 1294247 | 2062607 | 1453842 | 29\% |
| - | 1905201 | 0,97 | 1303037 | 2055168 | 1446433 | 30\% |
| - | 1902716 | 0,98 | 1311759 | 2047788 | 1439093 | 31\% |
| - | 1900236 | 0,99 | 1320412 | 2040467 | 1431819 | 32\% |
| - | 1897759 | 1,00 | 1328999 | 2033203 | 1424612 | 33\% |
| - | 1895286 | 1,01 | 1337519 | 2025996 | 1417470 | 34\% |
| - | 1892818 | 1,02 | 1345973 | 2018847 | 1410393 | 35\% |
| - | 1890354 | 1,03 | 1354362 | 2011753 | 1403379 | 35\% |
| - | 1887893 | 1,04 | 1362686 | 2004715 | 1396430 | 36\% |
| - | 1885437 | 1,05 | 1370946 | 1997731 | 1389543 | 37\% |


| - | 1882985 | 1,06 | 1379143 | 1990803 | 1382718 | $38 \%$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| - | 1880537 | 1,07 | 1387277 | 1983928 | 1375954 | $39 \%$ |
| - | 1878094 | 1,08 | 1395348 | 1977107 | 1369250 | $39 \%$ |
| - | 1875654 | 1,09 | 1403358 | 1970338 | 1362607 | $40 \%$ |

Table 8.8.3.3. NE Atlantic Mackerel. Short-term prediction: Management option table for 1001 kt catch in 2018 and a range of catch options in 2019.

| Rationale | Catch <br> (2019) | Fbar (2019) | $\begin{aligned} & \text { SSB } \\ & (2019) \end{aligned}$ | $\begin{aligned} & \text { SSB } \\ & (2020) \end{aligned}$ | $\begin{aligned} & \text { \% SSb } \\ & \text { change } \end{aligned}$ | \% catch <br> change | \% advice <br> change |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY AR | 318403 | 0,173 | 2117257 | 2333959 | 10\% | -68\% | -42\% |
| Catch(2019) = Zero | 0 | 0,000 | 2167164 | 2642785 | 22\% | -100\% | -100\% |
| $\begin{aligned} & \text { Catch }(2019)=2018 \text { catch } \\ & -20 \% \end{aligned}$ | 800447 | 0,498 | 2027332 | 1885650 | -7\% | -20\% | 45\% |
| Catch $(2019)=2018$ catch | 1000559 | 0,665 | 1983069 | 1707074 | -14\% | 0\% | 82\% |
| $\begin{aligned} & \text { Catch(2019) }=2018 \\ & +20 \% \end{aligned}$ | 1200671 | 0,858 | 1933279 | 1533339 | -21\% | 20\% | 118\% |
| $\operatorname{Fbar}(2019)=0.23$ | 413049 | 0,230 | 2101132 | 2244034 | 7\% | -59\% | -25\% |
| $\operatorname{Fbar}(2019)=0.31(\mathrm{Fpa})$ | 538181 | 0,310 | 2078758 | 2126532 | 2\% | -46\% | -2\% |
| $\operatorname{Fbar}(2019)=0.43$ (Flim) | 710360 | 0,430 | 2045756 | 1967543 | -4\% | -29\% | 29\% |
| $\operatorname{Fbar}(2019)=0.21$ (Fmsy) | 380375 | 0,210 | 2106773 | 2274977 | 8\% | -62\% | -31\% |
| $\operatorname{Fbar}(2019)=0.26$ | 460999 | 0,260 | 2092706 | 2198818 | 5\% | -54\% | -16\% |
| $\operatorname{Fbar}(2019)=0.27$ | 476705 | 0,270 | 2089907 | 2184058 | 5\% | -52\% | -13\% |
| $\operatorname{Fbar}(2019)=0.28$ | 492275 | 0,280 | 2087113 | 2169451 | 4\% | -51\% | -11\% |
| $\begin{aligned} & \text { SSB(2020) = MSY Btrig- } \\ & \text { ger = Bpa } \end{aligned}$ | 78048 | 0,040 | 2155496 | 2566214 | 19\% | -92\% | -86\% |
| SSB(2020) = Blim | 737375 | 0,450 | 2040320 | 1942890 | -5\% | -26\% | 34\% |
| $\operatorname{Fbar}(2019)=$ F2018 | 744410 | 0,455 | 2038892 | 1936484 | -5\% | -26\% | 35\% |
| $\mathrm{F}=0.20$ | 363822 | 0,200 | 2109600 | 2290695 | 9\% | -64\% | -34\% |
| $\mathrm{F}=0.21$ | 380375 | 0,210 | 2106773 | 2274977 | 8\% | -62\% | -31\% |
| $\mathrm{F}=0.22$ | 396784 | 0,220 | 2103950 | 2259424 | 7\% | -60\% | -28\% |
| $\mathrm{F}=0.23$ | 413049 | 0,230 | 2101132 | 2244034 | 7\% | -59\% | -25\% |
| $\mathrm{F}=0.24$ | 429172 | 0,240 | 2098319 | 2228804 | 6\% | -57\% | -22\% |
| $\mathrm{F}=0.25$ | 445155 | 0,250 | 2095510 | 2213732 | 6\% | -56\% | -19\% |
| $\mathrm{F}=0.26$ | 460999 | 0,260 | 2092706 | 2198818 | 5\% | -54\% | -16\% |
| $\mathrm{F}=0.27$ | 476705 | 0,270 | 2089907 | 2184058 | 5\% | -52\% | -13\% |
| $\mathrm{F}=0.28$ | 492275 | 0,280 | 2087113 | 2169451 | 4\% | -51\% | -11\% |
| $\mathrm{F}=0.29$ | 507710 | 0,290 | 2084323 | 2154996 | 3\% | -49\% | -8\% |

### 8.15 Figures



Figure 8.4.2.1. NE Atlantic Mackerel. Commercial catches in 2017, quarter 1.


Figure 8.4.2.2. NE Atlantic Mackerel. Commercial catches in 2017, quarter 2.


Figure 8.4.2.3. NE Atlantic Mackerel. Commercial catches in 2017, quarter 3.


Figure 8.4.2.4. NE Atlantic Mackerel. Commercial catches in 2017, quarter 4.


Figure 8.5.2.1. NE Atlantic mackerel. Weights-at-age in the catch.


Figure 8.5.2.2. NE Atlantic mackerel. Weights-at-age in the stock.


Figure 8.5.3.1. NE Atlantic mackerel. Proportion of mature fish at age.


Figure 8.6.1.2.1.: Mean egg production (stage $1 \mathrm{eggs} / \mathrm{m}^{2} / \mathrm{day}$ ) by half ICES rectangle for all Mackerel and Horse Mackerel Egg Surveys (MEGS) stations sampled in 2016. Egg production values are square root transformed. Crosses denote locations where sampling was undertaken but where no spawning was recorded. Area in yellow denotes the maximum geographical survey extent for the western survey area. Area/stations capturing $50 \%$ of spawning activity within that year are overlaid in red


Figure 8.6.1.2.2: Mackerel stage 1 egg counts $/ \mathrm{m}^{2} / \mathrm{day}$, May/June 2017, for all relevant surveys and all stations. The coloured squares correspond to the observed temperature recorded at 20 m depth during the plankton deployments. The 200,1000 and 2000 m contours are included for reference.


Figure 8.6.1.2.3: Mackerel stage 1 egg counts $/ \mathrm{m}^{2} /$ day 2018, for all surveys/stations sampled. The coloured squares represent the temperature in degrees Celsius at 20 m depth recorded during the plankton deployments.


Figure 8.6.1.2.4: Results of analysed Icelandic Ecosystem surveys in the Nordic Seas in July-August (IESSNS) station results, July 2018. The coloured squares represent the temperature in degrees Celsius at 20 m depth recorded during the plankton deployments.


Figure 8.6.2.1. Distributions of modelled squared catch rates of mackerel at approximately 3-9 months of age in first and fourth quarter demersal trawl surveys. Left) average rates for cohorts from 1998-2015; and Right) 2015 cohort. See Jansen et al. (2015) for details.


Figure 8.6.2.2. IBTS recruitment index derived from square root transformed CPUE. See Jansen et al. (2015) for details.


Figure 8.6.3.1. Mackerel catch rates from surface trawl hauls (circle size represents catch rate in $\mathrm{kg} / \mathrm{km}^{2}$ ) overlaid on mean catch rate per standardized rectangle ( $1^{\circ}$ lat. $\times 2^{\circ}$ lon.) from the IESSNS survey in 2017 (a) and in 2018 (b).


Figure 8.6.3.2. Estimated mackerel total stock biomass, with $90 \%$ CI, from the IESSNS for the years included in the assessment. North Sea is excluded from biomass index calculations in 2018.


Figure 8.6.3.3. Mackerel numbers by age from the IESSNS survey in 2018, excluding North Sea. Boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 500 replicates using the StoX software (http://www.imr.no/forskning/prosjekter/stox/nb-no).


Figure 8.6.3.4. Internal consistency of the mackerel abundance index from the IESSNS surveys including data from 2012 to 2018, excluding North Sea in 2018. Ages indicated by white numbers in grey diagonal cells. Statistically significant positive correlations ( $p<0.05$ ) are indicated by regression lines and red cells in upper left half. Corelation coefficients ( $\mathbf{r}$ ) are given in the lower right half.


Figure 8.6.4.1. Biomass (mill t) estimates of mackerel at ages 2+, $3+$ and $4+$ based on RFID tagging data and recaptures at year 1 after release (YearsOut=1). Estimates are scaled to the $10 \%$ survival used in SAM. Estimates for release year 2017 is only based on landings in quarter 1 2018. Note that the mortality happening over the year is not taken into account in the estimation.


Figure 8.6.4.2. Biomass (mill $\mathbf{t}$ ) estimates of Age 2+ mackerel based on RFID tagging data and recaptures at different numbers of years after release (YearsOut=1-6). Estimates are scaled to the $\mathbf{1 0 \%}$ survival used in SAM. Note that the mortality happening over the year is not taken into account in the estimation.


Figure 8.6.4.3. Biomass (mill t) estimates by age for the years 2011-2016 based on RFID tagging data and recaptures at different numbers of years after release (YearsOut=1-5). Estimates are scaled to $10 \%$ survival used in SAM. Note that the mortality happening over the year is not taken into account in the estimation.


Figure 8.6.4.4. Illustration of the change in distribution of catches and biomass scanned for tags over the time series of RFID-tagging data. A marked change happened from 2014 onwards, when Icelandic, Faroese and Scottish factories installed RFID antenna systems. The pictures are from a map websolution (www.smartfishmap.hi.no) where it is given an overview of tagging experiments, scanned catches and recaptures, where it is possible to filter by year and factory, and where there also is a list of recaptures. All ICES rectangles with info are clickable for more info.


Figure 8.6.4.5. Suggestion of a possible split into 4 areas/seasons with scanned catches and recaptures handled, by area. Note that this also would imply that SAM would have to include mortality happening over the year for the tagging data. At present it is not taken into account whether recaptures are coming in quarter Q1, Q3 or Q4.


Figure 8.6.4.6. Biomass (mill $t$ ) estimates of mackerel at ages $2+$ based on RFID tagging data and recaptures at year 1 after release (YearsOut=1), and based on recaptures in 4 different areas/seasons. Estimates are scaled to the $10 \%$ survival used in SAM. Estimates for release year 2017 is only based on landings in quarter 1 2018. Note that the mortality happening over the year is not taken into account in the estimation.


Figure 8.6.4.7. Biomass (mill t) estimates of mackerel by age in 2011-2015 based on RFID tagging data and recaptures at 1-2 years out after release (YearsOut=1-2), and based on recaptures in 4 different areas/seasons. Estimates are scaled to the $\mathbf{1 0 \%}$ survival used in SAM. Note that the mortality happening over the year is not taken into account in the estimation.


Figure 8.6.4.8. Left: Abundance index in billions individuals ages 4-12 per release years 2011-2016. Right: Year class trends in abundance (log scale) 2011-2016 from the index. The index is based on RFID tagging experiments 2011-2016, and data from scanned catches and recaptures in the two first years after a release year (yearsout=1-2).The index is already scaled down to the $\mathbf{1 0 \%}$ survival estimated by SAM (see Table 8.6.4.3 for data).


Figure 8.6.4.9. Biomass (mill t) estimates of mackerel by age (and total estimate) in 2016 based on RFID tagging off Ireland and Iceland and recaptures in 2017. Estimates are scaled to the $\mathbf{1 0 \%}$ survival used in SAM. Note that the mortality happening over the year is not taken into account in the estimation.


Figure 8.6.5.2.1. PELACUS 0318 mackerel density distribution. Polygons are drawn to encompass the backscattering energy, and polygon colour indicates the mean density expressed as tonnes per squared nautical mile (<1; 1-10; 10-25; 25-50; 50-100; and $>100$ ).


Figure 8.6.5.2.2: Mackerel abundance and biomass estimates by age group in ICES Divisions 8c. and 9.a during PELACUS 0318.


Figure 8.6.5.2.3: Mackerel subsurface egg distribution (no eggs/m ${ }^{3}$ ) as recorded by CUFES during PELACUS 0318.


Figure 8.7.2.1. NE Atlantic mackerel. Parameter estimates from the SAM model (and associated confidence intervals) for the WGWIDE 2018 update assessment.
Age 3

Figure 8.7.2.2 NE Atlantic mackerel. Estimated AR1 error correlation structure for the observations from the IESSNS survey age 3 to 11 .


Figure 8.7.2.3. NE Atlantic mackerel. Correlation between parameter estimates from the SAM model for the WGWIDE 2018 update assessment.


Figure 8.7.2.4. NE Atlantic mackerel. One Step Ahead Normalized residuals for the fit to the: (1) catch data (catch data prior to 2000 were not used to fit the model); (2) SSB estimates from egg survey; (3) recruitment index from the western Europe bottom trawl IBTS Q1 and Q4 surveys; and (4) abundance estimates at age from IESSNS survey. Blue circles indicate positive residuals (observation larger than predicted) and filled red circles indicate negative residuals.


Figure 8.7.2.5. NE Atlantic mackerel. One step ahead residuals for the fit to the recaptures of tags in the final assessment. The $x$-axis represents the release year, and the $y$-axis is the number of years between tagging and recapture. Each panel correspond to a given age at release. Blue circles indicate positive residuals (observation larger than predicted) and filled red circles indicate negative residuals.



Figure 8.7.2.6. NE Atlantic mackerel. Leave one out assessment runs. SAM estimates of SSB and Fbar, for assessments runs leaving out one of the observation data sets. 2018 WGWIDE assessment (black) and current assessment leaving out: egg survey (purple), the recruitment index (light blue), IESSNS index (seagreen) and without tagging data (dark green).


Figure 8.7.2.7. NE Atlantic mackerel. Leave one out assessment run excluding the RFID tagging data.Comparison of stock estimates from the 2018 WGWIDE assessment (blue) and the 2018 WGWIDE assessment without the 2017 RFID tagging data (red).


Figure 8.7.2.8. NE Atlantic mackerel. Influence of the latest year of data (recaptures from 2017) for the RFID tags on the output of the assessment. Comparison of stock estimates from the 2018 WGWIDE assessment (blue), the 2018 WGWIDE assessment without the 2017 recaptures (red) and the 2017 WGWIDE assessment (green).


Figure 8.7.2.9. NE Atlantic mackerel. Influence of the latest year of data for the IESSNS survey on the output of the assessment. Comparison of stock estimates from the 2018 WGWIDE assessment (blue) and the 2018 WGWIDE assessment without the 2018 IESSNS index (red).


Figure 8.7.3.1. NE Atlantic mackerel. Perception of the NEA mackerel stock, showing the SSB Fbar4-8 and recruitment (with $95 \%$ confidence intervals) from the SAM assessment.


Figure 8.7.3.2. NE Atlantic mackerel. Estimated selectivity for the period 1990 to 2017, calculated as the ratio of the estimated fishing mortality-at-age and the Fbart-8 value in the corresponding year.


Figure 8.7.4.1.1. NE Atlantic mackerel. Comparison of estimated model parameters for the WGWIDE 2018 update assessment (blue), the assessment run with the same configuration on the RFID tag dataset structured by recapture area (green), and the assessment with survival rate for the RFID tag estimated for each area (red).


Figure 8.7.4.1.2. NE Atlantic mackerel. Comparison of estimated post release survival rates for the WGWIDE 2018 update assessment (blue), the assessment run with the same configuration on the RFID tag dataset structured by recapture area (green), and the assessment with survival rate for the RFID tag estimated for each area (red).


Figure 8.7.4.1.3. NE Atlantic mackerel. Comparison of the stock trajectories between the WGWIDE 2018 update assessment (blue), the assessment run with the same configuration on the RFID tag dataset structured by recapture area (red), and the assessment with survival rate for the RFID tag estimated for each area (green).


Figure 8.7.4.2.1. NE Atlantic mackerel. Residuals (OAS) for the RFID tags grouped by year-class and age at release and centred. The different panels correspond to different year-classes. Green circles indicate positive residuals (observation larger than predicted) and filled red circles indicate negative residuals.


Figure 8.7.4.2.2. NE Atlantic mackerel. Comparison of the stock trajectories between the WGWIDE 2018 update assessment (black) and the same assessment using only the RFID tag data corresponding to the first 2 years of recapture after tagging (red).


Figure 8.7.5.1. NE Atlantic mackerel. Uncertainty (standard deviation of the log values) of the estimates of SSB and Fbar from the SAM for the 2018 WGWIDE assessment.


Figure 8.7.5.2. NE Atlantic mackerel. Analytical retrospective patterns (5 years back) of SSB, Fbar48 and recruitment from the WGWIDE 2018 update assessment.


Figure 8.7.5.3. NE Atlantic mackerel. Process error expressed as annual deviations of abundances at age, for the 2018 WGWIDE assessment (orange) and from the 2017 WGWIDE assessment (black).

Process error deviation in biomass


Figure 8.7.5.4. NE Atlantic mackerel. Model process error expressed in biomass cumulated across age-group for the 2018 WGWIDE assessment.


Figure 8.10.1. NE Atlantic mackerel. Comparison of the stock trajectories between the 2018 WGWIDE assessment (blue)and the 2017 assessment (red).


Figure 8.10.2. NE Atlantic mackerel. Comparison of model parameters and their uncertainty for the 2018 WGWIDE (green) and the 2017 WGWIDE assessment (red).


Figure 8.10.3. NE Atlantic mackerel. Comparison of the joint uncertainty on recent estimates of SSB and Fbar for the WGWIDE 2018 update assessment and last year's assessment.

## 9 Red gurnard in the Northeast Atlantic

### 9.1 General biology

The main biological features known for red gurnard (Aspitrigla (Chelidonichthys) cucu${ }^{l}$ lus) are described in the stock annex. This species is widely distributed in the Northeast Atlantic from South Norway and North of the British Isles to Mauritania on grounds between 20 and 250 m . This benthic species is abundant in the Channel (7de) and on the shelf West of Brittany (7h, 8a), living on gravel or coarse sand. In the Channel, the size at first maturity is $\sim 25 \mathrm{~cm}$ at 3 years old (Dorel, 1986).

### 9.2 Stock identity and possible assessments areas

A compilation of datasets from bottom-trawl surveys undertaken within the project 'Atlas of the marine fishes of the northern European shelf' has produced a distribution map of red gurnard. Higher occurrences of red gurnard with patchy distribution have been observed along the Western approaches from the Shetlands Islands to the Celtic Seas and the Channel.

A continuous distribution of fish crossing the Channel and the area West of Brittany does not suggest a separation of the Divisions 7 d from 7 e and 7 h . Therefore a split of the population between the Ecoregions does not seem appropriate. Similar temporal signals observed in NS-IBTS and SCO-WCIBTS surveys, which are not seen in other survey series, may suggest a linkage between subareas 4 and 6 . Further investigations are needed to progress on stocks boundaries such as morphometric studies, tagging and genetic population studies.

### 9.3 Management regulations

There is currently no technical measure specifically applied to red gurnard or other gurnard species. The exploitation of red gurnard is submitted to the general regulation in the areas where they are caught. There is no minimum landing size set.

### 9.4 Fisheries data

Red gurnard is mainly landed as bycatch by demersal trawlers in mixed fisheries, predominantly in Divisions 7d, 7e and 7h (Figure 9.1). High discard rates and lack of resolution at a species level make interpretation of spatial trends in catches in other areas problematic.

### 9.4.1 Historical landings

Official landings reported at ICES are available in Table 9.1 and Table 9.2. Before 1977, red gurnard was not specifically reported. Landings of gurnards are still not always reported at a species level, but rather as mixed gurnards. This makes interpretations of the records of official landings difficult.
International landings have fluctuated between 3452-5171 tonnes since 2006. France is the main contributor of 'red gurnard' landings, with around $80 \%$ of landings coming from ICES Subarea 7d-h (Celtic Sea/English Channel). In the North Sea red gurnard landings are variable, but roughly evenly distributed between Divisions 4a,b and c. Landings from the west of Scotland and Ireland, and the Irish Sea (ICES Subarea 6a-b, 7a-c, 7j) and Bay of Biscay (ICES Division 8) have been consistently low.

### 9.4.2 Discards

Discard data for red gurnard has been provided for 2015, 2016 and 2017 through Intercatch (Table 9.3). For those countries which provided data, discard rates ranged between from $48 \%$ and $91 \%$ of catch in 2017 (Table 9.4).

### 9.5 Survey data

Information on gurnard abundance are available in DATRAS for the IBTS-Q1 survey in the North Sea, Scottish West Coast Groundfish Survey (WCGFS), Irish Groundfish Survey (IGFS) and the French EVHOE-WIBTS-Q4 survey in the Celtic Sea and Bay of Biscay and CGFS-Q4 in Division 7d. Each of these surveys covers a specific area of red gurnard distribution. Lengths at age are available from CGFS-Q4 in and IGFS-Q4

- • NS- IBTS-Q1 series. Before 1990, red gurnard was scarce in North Sea and the abundance index was close to 0 . The abundance index of red gurnard has trended generally upwards between 1994-2013, before declining somewhat, although it remains well above long-term average values. This change reflects an increase of the abundance in the northern and central North Sea (4a-b). It is interesting to contrast these trends with the apparent very low abundances in the NS-IBTS-Q3 series.
-     - SCO-WCGFS series. Before 1996, red gurnard was also scarce on the west of Scotland. The abundance index trended strongly upwards after 1997, reaching a peak in 2013, before declining to around the series average in recent years.
-     - IGFS series. The abundance index of red gurnard in the IGFS series has varied around the series mean without trend between 2002 and 2017.
-     - CGFS-Q4 series. Over the time-series 1988-2011, the abundance index has fluctuated, peaked in 1994, reached a low in 2011, but is above long term mean in 2016.
- • EVHOE-WIBTS-Q4 series. Over the period 1997-2011, the abundance index in Nb or $\mathrm{kg} / \mathrm{hr}$ has increased over time. Age reading of red gurnards caught during EVHOE survey has been carried out in 2006 and routinely since 2008. They indicate that the individuals caught are mainly of age 1 and 2.

Survey abundance information was provided separately for the Spanish Porcupine and Northern Spanish groundfish surveys (SP-PORC and SP-NSGFS). Both survey indices are variable, but show an overall upwards trend over time in numbers and weight per tow.

### 9.6 Biological sampling

Number at length information was provided by French and Portuguese landings and discards. There remains a lack of regular sampling for red gurnard in commercial landings and discarding to provide series of length or age compositions usable for a preliminary analytical assessment.

### 9.7 Biological parameters and other research

There is no update of growth parameters and available parameters from several authors are summarized in the Stock Annex. They vary widely. Available length-weight relationships are also shown in Stock Annex. Natural mortality has not been estimated in the areas studied at this Working Group.

### 9.8 Analyses of stock trends

In the North Sea, the appearance of red gurnard in the index of the IBTS Survey since 1990 is in line with an increase of the abundance in 4a. In Eastern Channel, the abundance index of the CGFS-Q4 survey has widely fluctuated, with a weak decline. The EVHOE-WIBTS-Q4 survey has slightly increased since its beginning in the 1990s.

### 9.9 Data requirements

Gurnards are still not always reported by species, but rather as mixed gurnards (see WDocument08 and WDocument09). This makes interpretations of the records of official landings difficult. Extending the studied area by a survey in 7 e and collecting length and age data of red gurnard in the main area of production should help in better understanding the biology and dynamics of this species.

### 9.10 References

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### 9.11 Tables

Table 9.1 Red gurnard in the Northeast Atlantic official landings by country in tonnes.

| Year | Belgium | Spain | France | Jersey | Guernsey | IRELAND | IM | Netherlands | Portugal | UK | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 313 | 0 | 4552 | 10 | 0 | 0 | 0 | 57 | 125 | 115 | 5172 |
| 2007 | 328 | 0 | 4494 | 4 | 0 | 0 | 1 | 66 | 127 | 156 | 5176 |
| 2008 | 352 | 0 | 4045 | 8 | 0 | 0 | 0 | 92 | 112 | 166 | 4776 |
| 2009 | 227 | 0 | 3310 | 6 | 0 | 1 | 0 | 160 | 150 | 263 | 4118 |
| 2010 | 237 | 0 | 3437 | 2 | 0 | 0 | 0 | 251 | 115 | 362 | 4403 |
| 2011 | 306 | 0 | 3176 | 2 | 0 | 1 | 1 | 295 | 134 | 257 | 4172 |
| 2012 | 306 | 0 | 2706 | 4 | 26 | 0 | 3 | 329 | 148 | 257 | 3778 |
| 2013 | 288 | 576 | 3154 | 9 | 16 | 2 | 3 | 267 | 113 | 329 | 4756 |
| 2014 | 263 | 399 | 3782 | 6 | 0 | 5 | 3 | 241 | 108 | 283 | 5089 |
| 2015 | 187 | 91 | 2919 | 3 | 0 | 0 | 2 | 210 | 122 | 341 | 3874 |
| 2016 | 238 | 87 | 2598 | 0 | 9 | 1 | 3 | 224 | 106 | 381 | 3646 |
| 2017* | 265 | 103 | 2396 | 0 | 0 | 9 | 1 | 226 | 126 | 327 | 3452 |
| 2017** | 258 | 61 | 2410 |  |  | 9 |  | 228 |  | 94 | 3061 |

*Preliminary Data
**Intercatch Data

## Table 9.2 Red gurnard in the Northeast Atlantic official landings by area in tonnes.

*Preliminary Data,

| Year | 4A | 4B | 4c | 5B | 6A | 6B | 7A | 7в | 7c | 7D | 7E | 7F | 7G | 7H | 7J | 7NK | 8A | 8B | 8 C | 8D | 9A | 9NK | 10A | 10NK | 14A | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 13 | 83 | 64 | 0 | 32 | 1 | 11 | 9 | 12 | 1101 | 2803 | 229 | 16 | 446 | 5 | 1 | 153 | 60 | 1 | 5 | 9 | 115 | 0 | 1 | 0 | 5171 |
| 2007 | 12 | 120 | 55 | 2 | 21 | 0 | 7 | 7 | 15 | 1229 | 2674 | 246 | 15 | 437 | 4 | 0 | 139 | 59 | 3 | 2 | 125 | 0 | 0 | 2 | 0 | 5175 |
| 2008 | 34 | 64 | 54 | 0 | 28 | 3 | 5 | 7 | 16 | 1236 | 2451 | 249 | 9 | 408 | 5 | 0 | 66 | 24 | 3 | 1 | 109 | 0 | 3 | 0 | 0 | 4775 |
| 2009 | 58 | 59 | 92 | 0 | 94 | 2 | 4 | 8 | 6 | 1293 | 1557 | 112 | 22 | 510 | 7 | 0 | 98 | 40 | 1 | 3 | 148 | 0 | 1 | 0 | 0 | 4115 |
| 2010 | 79 | 63 | 86 | 0 | 101 | 46 | 13 | 8 | 10 | 1531 | 1608 | 132 | 23 | 433 | 9 | 0 | 100 | 33 | 0 | 2 | 114 | 0 | 0 | 1 | 0 | 4393 |
| 2011 | 66 | 29 | 51 | 0 | 69 | 54 | 13 | 5 | 6 | 1295 | 1753 | 124 | 20 | 372 | 9 | 0 | 112 | 46 | 1 | 3 | 133 | 0 | 1 | 0 | 1 | 4164 |
| 2012 | 83 | 71 | 78 | 0 | 51 | 7 | 8 | 2 | 5 | 1244 | 1441 | 145 | 53 | 294 | 2 | 0 | 83 | 50 | 8 | 1 | 136 | 4 | 1 | 0 | 1 | 3770 |
| 2013 | 88 | 109 | 60 | 0 | 47 | 0 | 10 | 2 | 6 | 1193 | 1692 | 170 | 58 | 477 | 2 | 0 | 79 | 72 | 532 | 1 | 155 | 0 | 2 | 0 | 0 | 4755 |
| 2014 | 102 | 52 | 68 | 0 | 47 | 3 | 7 | 1 | 2 | 1294 | 1642 | 115 | 19 | 1069 | 1 | 0 | 82 | 75 | 363 | 3 | 139 | 0 | 3 | 0 | 0 | 5088 |
| 2015 | 133 | 102 | 53 | 0 | 58 | 1 | 4 | 3 | 1 | 790 | 1553 | 87 | 6 | 703 | 1 | 0 | 95 | 70 | 81 | 2 | 128 | 0 | 2 | 0 | 0 | 3874 |
| 2016 | 112 | 83 | 117 | 0 | 76 | 1 | 11 | 3 | 1 | 906 | 1268 | 114 | 16 | 608 | 1 | 0 | 87 | 63 | 56 | 1 | 120 | 0 | 1 | 0 | 0 | 3645 |
| 2017* | 52 | 43 | 87 | 0 | 29 | 1 | 11 | 0 | 0 | 868 | 1424 | 83 | 38 | 473 | 3 |  | 77 | 48 | 58 | 1 | 154 |  | 1 |  |  | 3453 |

Table 9.3 Red gurnard in the Northeast Atlantic, discards (t) by country, 2015-2017.

|  | COUNTRY | 2015 | 2016 |
| :--- | :---: | :---: | :---: |
| France | 1323 | 2249 | 2017 |
| Ireland | 10 | 147 | 93 |
| Spain |  | 286 | 272 |
| UK (ENG) | 74 | 30 | 198 |
| UK (SCO) | 649 | 411 | 2796 |
| Total | 2057 | 3125 |  |

Table 9.4. Discarding of Red gurnard in the Northeast Atlantic, as a percentage of catch, by country, in 2017.

|  | COUNTRY |
| :--- | :---: |
| France | DISCARD RATE (\%) |
| Ireland | 48 |
| Spain | 91 |
| UK (SCO) | 82 |

### 9.12 Figures



Figure 9.1.Red gurnard in the Northeast Atlantic. Landings in 2017, by statistical rectangle, from BEL, FRA, UK(E\&W), UK(IoM) \& UK(SCO).

## 10 Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8, and 9a

### 10.1 General biology

Striped red mullet (Mullus surmuletus) is a predominantly benthic species found along the coasts of Europe, southern Norway and northern Scotland (northern Atlantic, Baltic Sea, North Sea and the English Channel), up to the Northern part of West Africa, in the Mediterranean Basin, and in the Black Sea (Hureau, 1986; Mahé et al., 2005). Young fish are distributed in lower salinity coastal areas, while adults have a more offshore distribution.

Adult red mullet feed on small crustaceans, annelid worms and molluscs, using their chin barbels to detect prey and search the mud. As a consequence, striped red mullet are typically found on sandy, gravelly and shelly sediments where they can excavate sediment with their barbels and dislodge the small invertebrates. The main natural predators of striped red mullet are sea basses, pollacks, barracudas, monkfish, congers and sharks (Caill-Milly et al., 2017).

Sexual maturity is reached at the beginning of the second year for males, followed by a marked decrease in growth rates, and at the end of the second or beginning of the third year for females which therefore continue their rapid growth a little longer (Déniel, 1991). In the English Channel, this species matures at approximately 16 cm (Mahé et al., 2005), while in the Bay of Biscay, the sizes of first sexual maturity are given by Dorel (1986) as: males 16 cm , females 18 cm and a length at which $50 \%$ of the individuals are mature (the distinction between the two sexes is not mentioned) of 22 cm .

Spawning occurs in the spring and early summer (May to June according to Desbrosses, 1935) with a spawning peak in June in the northern Bay of Biscay (N'Da \& Déniel, 1993). Eggs and larvae average 2.8 mm and are pelagic (Jones, 1972; Russell, 1976). The hatching takes place after three days at $18^{\circ} \mathrm{C}$ and after eight days at a temperature of $9^{\circ} \mathrm{C}$ (Quéro \& Vayne, 1997). After metamorphosis juveniles become first demersal then benthic. At the age of one month, they measure about 5 cm and weigh 0.9 to 1.6 g . They show rapid growth during their first four months of life between July and October. Increases in length and mass are about 7 cm and 25 g on average during this period (N'Da \& Déniel, 2005). The rate of growth declines sharply in October due to the cooling of water and the scarcity of trophic resources in the environment. These conditions contribute to the initiation of migration of red mullets to greater depths offshore. Until the age of two, there is no significant difference in size between males and females; they then measure $20-23 \mathrm{~cm}$. Sexual dimorphism is observed from the age of first maturity due to growth rates that will then differ between the two sexes. From age three, females exceed males in length by 4 cm on average and 7 cm beyond 5 years (N'Da \& Déniel, 2006).

The maximum reported age of the striped red mullet is 11 years (Quéro \& Vayne, 1997; ICES, 2012), while the maximum length given is 44.5 cm in the Bay of Biscay (Dorel, 1986) and 40 cm elsewhere (Hureau, 1986; Bauchot, 1987). The maximum reported mass is 1 kg (Muus and Nielsen, 1999).

### 10.2 Management regulations

Prior to 2002, France enforced a minimum landing size of 16 cm . Since this minimal size requirement has been removed, immature individuals $(<14 \mathrm{~cm})$ have been recorded in landings. There is no TAC for this stock.

### 10.3 Stock ID and possible management areas

In 2004 and 2005, a study using fish geometrical morphometry was carried out in the Eastern English Channel and the Bay of Biscay. It pointed out a morphological difference on striped red mullets between those from the Eastern English Channel and those from the Bay of Biscay.

Benzinou et al. (2013) conducted stock identification studies based on otolith and fish shape in European waters and showed that striped red mullet can be geographically divided into three zones:

-     - The Bay of Biscay (Northern Bay of Biscay - NBB, and Southern Bay of Biscay - SBB)
- A mixing zone composed of the Celtic Sea and the Western English Channel (CS + WEC)
-     - A northern zone composed of the Eastern English Channel and the North Sea (EEC + NS)

The distinction between the putative Biscay and Western Channel/Celtic Sea populations is supported by the distribution of landings at a statistical rectangle level (Fig. 10.1). This assessment treats these putative components as one population. At present there are no management measures in place, however this structuring should be taken into account if measures are considered.

### 10.4 Fisheries data

Official landings have been recorded since 1975 and after early increases they have declined in recent years (Table 10.1). Landings are mainly taken from Subarea 7 and 8 (Table 10.2) and France accounts for the majority of removals. The striped red mullet is one species among set of benthic (demersal) species targeted by the French fleet, and is mainly caught by bottom trawlers with a mesh size of $70-99 \mathrm{~mm}$. In the Western English Channel striped red mullet is also caught by gillnets. Danish seine appeared in 2008 as a result of some trawlers converting to use seine gears.

The average characteristics of vessels in French fleets that caught red mullet from 2000 to 2015 are: 41.1 GRT, 191.1 kW engine power, 12.9 m length and 22 years of service. Net vessels are made up of the smallest units ( $85 \%$ are less than 12 m long), while $52 \%$ of bottom trawlers are less than 15 m ; the seiners are by far the largest and the oldest vessels (Caill-Milly et al., 2017).

The French activity on this species differs between the area composed by West Scotland/Celtic sea (including West Channel) and the area comprising the Bay of Biscay. In the first one, landings are mainly taken by bottom trawlers, followed by gillnet. In the second one, they are mainly done by bottom trawls, seine and nets. French activity in the Atlantic Iberian waters remains limited. The Spanish activity is located in the north (8.a,b) and the south (8.c) of the Bay of Biscay.

Prior to 2015 this species was not recorded as being discarded by French or Portuguese vessels and was infrequent in Spanish sampling. Discarding represented between 9\% and $68 \%$ of UK catches in 2014-17 (Table 10.3), however there are concerns about how
these discards have been estimated - the 2016 figure is based on a sample of 2 fishes. French discard estimates for 2017 represented 7\% of catch. For French demersal trawls ( $70-99 \mathrm{~mm}$ mesh size), discards are essentially composed of individuals measuring between 8 and 17 cm (Fig. 10.2).

### 10.5 Survey data, recruit series

Exchange data is available in Datras since 1997 for the French EVHOE survey, covering the Bay of Biscay and Celtic Sea, and from 2002 onwards for the Portuguese groundfish survey (PT-IBTS), covering the Portuguese coast. Standardised catch rates in the EVHOE survey are variable around the series mean between 1997-2011, before falling to a lower level thereafter. Similarly, catch rates in the PT-IBTS are at a low level in 2005, peak in 2010, before falling back to near the series mean in recent years (Fig. 10.3).

Abundance indices per size class during EVHOE-WIBTS-Q4 show mainly fish between 8 and 17 cm (TL).

Data was provided separately for the northern Spanish groundfish survey (SP-NSGFS), showing a similar variable trend to the EVHOE survey in the early part of the series, followed by a decline to lower levels in recent years (Fig. 10.4).

### 10.6 Biological sampling

In the Bay of Biscay sexual maturity and length measures were taken in 2009 by AZTI. French samplings started in 2004 in the Eastern Channel and in the south North Sea, and since 2008 in the Bay of Biscay.

### 10.7 Biological parameters and other research

Since 2004, data (age, length, sexual maturity) are usually collected by France for the Eastern English Channel and the southern North Sea. France started to collect data for $8 \mathrm{a}, \mathrm{b}$ at the end of 2007. In 2007-2008, the striped red mullet otolith exchange had for goal to optimize age estimation between countries.
In 2011, an Otolith Exchange Scheme was carried out, which was the second exercise for the Striped red mullet (Mullus surmuletus). Four readers of this exchange interpreted an images collection coming from the Bay of Biscay, the Spanish coasts and the Mediterranean coasts (Spain and Italy). A set of Mullus surmuletus otoliths (N=75) from the Bay of Biscay presented highest percentage of agreement ( $82 \%$ ). On 75 otoliths, 34 were read with $100 \%$ agreement $(45 \%)$ and thus a CV of $0 \%$. Modal age of these fishes was comprised between 0 and 3 years (Mahé et al., 2012).

### 10.8 Analysis of stock trends/ assessment

Currently, age structured analytical stock assessment is not possible due to a too short time-series of available data.

### 10.9 Data requirements

Regular sampling of biological parameters of striped red mullet catches must be continued under DCF. Sampling in the Celtic Sea and in the Bay of Biscay started in 2008. In 2010 and 2011, sampling for age and maturity data was reduced compared to 2009, due to the end of the Nespman project. Since 2009, a concurrent sampling design carried out, should provide more data (length compositions) than in recent years.

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### 10.11 Tables

Table 10.1 Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8, and 9a official landings by country in tonnes

| Year | Belgiu <br> M | $\begin{gathered} \text { SPAI } \\ \mathrm{N} \end{gathered}$ | Fran CE | Guerns EY | IreLA ND | Jersey | Netherla NDS | Portu GAL | UK | Tот AL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 |  |  |  |  |  |  |  |  | 17 | 267 |
|  | 33 | 379 | 1944 | 8 | 15 | 1 | 115 | 11 | 0 | 5 |
| 2007 |  |  |  |  |  |  |  |  | 19 | 294 |
|  | 43 | 390 | 1926 | 9 | 17 | 1 | 148 | 222 | 3 | 9 |
| 2008 |  |  |  |  |  |  |  |  | 16 | 231 |
|  | 26 | 379 | 1384 | 9 | 17 | 0 | 165 | 169 | 4 | 4 |
| 2009 |  |  |  |  |  |  |  |  | 13 | 250 |
|  | 20 | 490 | 1539 | 5 | 10 | 0 | 110 | 199 | 1 | 4 |
| 2010 |  |  |  |  |  |  |  |  | 13 | 275 |
|  | 20 | 465 | 1725 | 5 | 5 | 0 | 128 | 276 | 2 | 6 |
| 2011 |  |  |  |  |  |  |  |  | 15 | 278 |
|  | 21 | 504 | 1722 | 0 | 5 | 0 | 130 | 245 | 4 | 2 |
| 2012 |  |  |  |  |  |  |  |  | 12 | 215 |
|  | 37 | 328 | 1318 | 0 | 4 | 1 | 125 | 217 | 2 | 2 |
| 2013 |  |  |  |  |  |  |  |  |  | 151 |
|  | 28 | 245 | 925 | 5 | 3 | 0 | 50 | 187 | 70 | 4 |
| 2014 |  |  |  |  |  |  |  |  |  | 147 |
|  | 12 | 265 | 914 | 5 | 2 | 0 | 1 | 221 | 53 | 4 |
| 2015 |  |  |  |  |  |  |  |  | 10 | 198 |
|  | 23 | 248 | 1207 | 5 | 3 | 0 | 110 | 282 | 2 | 0 |
| 2016 |  |  |  |  |  |  |  |  |  | 174 |
|  | 28 | 194 | 1166 | 0 | 4 | 0 | 69 | 204 | 83 | 8 |
| 2017* | 35 | 118 | 988 | 5 | 10 |  | 16 | 157 | 64 | 139 |
|  |  |  |  |  |  |  |  |  |  | 3 |
| $2017^{*}$ | 36 | 328 | 997 | 0 | 10 |  | 13 | 154 | 64 | $\begin{gathered} 160 \\ 2 \end{gathered}$ |

* Preliminary Data
** Intercatch Data

Table 10.2 Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8, and 9a official landings by area in tonnes

| Year | 6A | 6B | 7A | 7B | 7c | 7E | 7F | 7G | 7H | 7J | 7k | 8A | 8B | 8c | 8D | 8E | 9A | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 0 | 0 | 1 | 1 | 0 | 869 | 50 | 24 | 103 | 5 | 0 | 1023 | 468 | 71 | 14 | 0 | 39 | 2668 |
| 2007 | 1 | 0 | 1 | 1 | 1 | 1047 | 54 | 22 | 104 | 12 | 0 | 861 | 473 | 90 | 16 | 0 | 267 | 2949 |
| 2008 | 0 | 0 | 1 | 1 | 0 | 880 | 46 | 16 | 73 | 13 | 0 | 639 | 246 | 87 | 18 | 0 | 296 | 2314 |
| 2009 | 2 | 0 | 1 | 2 | 1 | 592 | 25 | 9 | 74 | 17 | 0 | 879 | 460 | 156 | 44 | 0 | 243 | 2504 |
| 2010 | 2 | 0 | 1 | 3 | 1 | 642 | 26 | 10 | 59 | 16 | 1 | 1033 | 467 | 146 | 19 | 0 | 331 | 2756 |
| 2011 | 1 | 1 | 1 | 0 | 0 | 665 | 20 | 10 | 55 | 6 | 0 | 970 | 513 | 214 | 17 | 0 | 310 | 2782 |
| 2012 | 0 | 0 | 0 | 0 | 0 | 493 | 23 | 7 | 34 | 4 | 0 | 696 | 387 | 200 | 27 | 0 | 280 | 2152 |
| 2013 | 0 | 0 | 0 | 1 | 0 | 232 | 23 | 7 | 36 | 2 | 0 | 473 | 328 | 166 | 6 | 0 | 241 | 1514 |
| 2014 | 1 | 0 | 0 | 0 | 0 | 192 | 15 | 3 | 40 | 1 | 0 | 523 | 240 | 151 | 12 | 0 | 297 | 1474 |
| 2015 | 0 | 0 | 0 | 1 | 0 | 595 | 10 | 2 | 35 | 1 | 0 | 506 | 327 | 127 | 7 | 0 | 369 | 1980 |
| 2016 | 0 | 0 | 0 | 2 | 0 | 417 | 21 | 7 | 35 | 3 | 0 | 549 | 311 | 117 | 10 | 0 | 277 | 1748 |
| 2017* | 0 | 0 | 0 | 1 | 0 | 283 | 26 | 21 | 36 | 0 | 0 | 505 | 244 | 82 | 5 | 0 | 185 | 1393 |
| 2017** | 0 | 0 | 0 | 1 | 0 | 277 | 27 | 21 | 37 | 3 | 0 | 514 | 324 | 160 | 5 | 0 | 231 | 1601 |

* Preliminary Data
** Intercatch Data

Table 10.3 Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8, and 9a discards (t) by country in 2012-2016

| COUNTRY | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| BE |  |  |  |  | 2 | 0 |
| ES |  | 4 | 5 | 8 | 74 |  |
| FR |  |  | 115 | 213 | 0 |  |
| IE | 0.0 | 0.0 | 0.0 |  | 0 | 0 |
| PT | 2 | 1 | 5 | 77 | 171 | 11 |
| UK | 2 | 1 | 9 | 197 | 392 | 87 |
| Total |  |  |  |  |  |  |

### 10.12 Figures



Figure 10.1. Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Landings by statistical rectangle in 2017 for BEL, FRA, IRE, UK (E\&W) \& UK (SCO).


Fig 10.2. Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Length distribution of French catches from OTB_DEF_70-99.


Figure 10.3. Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Standardised survey abundances for EVHOE (1997-2017) and Portuguese IBTS (2002-2017) surveys.


Figure 10.4. Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Standardised survey abundances for SP-NSGFS (1983-2017).

## Annex 1: List of Participants

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28 August- 3 September 2018

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## Annex 2: Recommendations

| Recommendation | To |
| :---: | :---: |
| It is recommended that a new Working Group will be established to address methodology and dataanalyses in relation tothe RFID taggings of mackerel (and NSS-herring) <br> There is a need for international body (e.g. WG comparable to survey groups) to analyse and explore the recapture data etc. <br> A proposal of ToRs for this WG can be found in the Report section 8.6.4.. | ACOM |
| It is recommended that differing national approaches to the assignment of mixed gurnard catches to species level be reviewed in order to develop a standardised procedure which can be used going forwards, and investigate the assignment of historical mixed catches. <br> Large catches of mixed gurnards (GUX) are still reported from several countries. This has a strong, negative impact on the development of future assessments and advice of the stock. <br> Section in the report this relates to: 9.9. | WGCATCH |
| It is recommended that any available data on stock structuring of red gurnards in Divs. 3-8 (and elsewhere) are presented to the ICES Stock Identification Methods Working Group for future consideration on the stock identity and structure of this species before a new benchmark workshop is considered Based on differing trends in survey abundances there appears to be spatial structuring of red gurnard populations within the area considered in the present assessment unit (Div. 3-8). In order to further develop the assessment of this species it is important that stock structuring is taken into account | SIMWG |
| WGWIDE recommends that IBWSS explores methods/approaches to survey division 8abd in order to understand the dynamics and connectivity between blue whiting spawning components <br> IBWSS covers the core spawning area of blue whiting, but little is known about the connectivity between this area and the possible southern spawning areas as revealed in recent research papers. <br> Section in the report this relates to: 2.3.7.2 | WGIPS |
| It is recommended that work is initiated on how to separate among different stock components of herring in internationally coordinated surveys <br> In the IESNS survey other herring stocks (e.g. Icelandic summer spawning herring and North Sea herring) are found in the boundary regions of the survey area that may mix with the NSS herring in the survey area. <br> Section in the report this relates to: 4.14 | WGIPS, WGBIOP |
| It is recommended to age read mackerel in the EVHOE survey starting from Q42018 <br> Catch rates of age 0 mackerel from the EVHOE survey are used for the recruitment index. The age 0 mackerel are currently separated from age 1 mackerel by length frequency distributions, because the mackerel are not aged. Section in the report this relates to: 8.6.2 | IBTSWG |
| Increase the spatial coverage of the IBTS in Q4 to include the areas that was covered by the English Q4SWIBTS survey up to 2011 <br> Since 2011, the English Q4SWIBTS survey (covering the Irish sea and the central-eastern part of the Celtic sea including the area around Cornwall) has been discontinued. In some years, this has been an important nursery area for mackerel and it is not completely covered by other surveys (Irish and French). Section in the report this relates to: 8.6.2 | IBTSWG |

Increase the spatial coverage of NS-IBTS Q1 to include the south-western
IBTSWG
Norwegian shelf and shelf edge in proximity to the Norwegian trench
The IBTS has observed high catch rates in some years at the north-eastern edge of the survey area (towards the Norwegian trench) in winter. It is therefore possible that some recruits are also overwintering on the other side of the trench along the south western shelf edge of Norway.
Section in the report this relates to: 8.6.2

## Annex 3: WGWIDE 2019 Terms of Reference

WGWIDE- Working Group on Widely Distributed Stocks
2018/2/ACOM23 The Working Group on Widely Distributed Stocks (WGWIDE), chaired by Gudmundur J. Óskarsson, Iceland, will meet in Spain (location tbd), 28 Au -gust-3 September 2019 to:
a) Address generic ToRs for Regional and Species Working Groups.
b ) Estimate MSY proxy reference points for the category 3 and 4 stocks in need of new advice in 2019:
i) Update the MSY proxy reference points for those category 3 and 4 stocks with existing proxy reference points using most recent data. For those stocks without reference points listed below, collate necessary data and information in order to estimate MSY proxy reference points prior to the Expert Group meeting. The official ICES data call included a call for length and life history parameters for each stock in the table below;
ii ) Propose appropriate MSY proxies for each of the stocks listed below by using methods provided in the ICES Technical Guidelines (ICES, 2017) along with available data and expert judgement.

| Stock Code | Stock name description | EG | Data Cate- <br> gory |
| :--- | :--- | :--- | :---: |
| gur.27.3-8 | Red gurnard (Chelidonichthys cuculus) in <br> subareas 3-8 (Northeast Atlantic) | WGWIDE | 6.2 |
| hom.27.3a4bc7d | Horse mackerel (Trachurus trachurus) in <br> divisions 3.a, 4.b-c, and 7.d (Skagerrak <br> and Kattegat, southern and central <br> North Sea, eastern English Channel) | WGWIDE | 3 |

The assessments will be carried out on the basis of the stock annex. The assessments must be available for audit on the first day of the meeting.

Material and data relevant for the meeting must be available to the group no later than 14 days prior to the starting date.

WGWIDE will report by 10 September 2019 for the attention of ACOM.

## Annex 4: List of Stock Annexes

| STOCK ID | STOCK NAME | LAST UPDATED |
| :--- | :--- | :--- |
| Loc.27.6-8 | Boarfish (Capros aper) in Sub areas 6-8 (Celtic Seas, English Channel, and Bay of <br> Biscay) |  |
|  | Herring (Clupea harengus) in subareas 1, 2, and 5, and in divisions 4.a and 14.a, <br> Norwegian spring-spawning herring (the Northeast Atlantic and Arctic Ocean) | 4 March 2016 |

## Annex 5: Audit Reports

# Audit of North Sea horse mackerel (WGWIDE 2018 ) 

Date: 11.09.2018
Auditor: Leif Nøttestad

## General

In 2012 the North Sea horse mackerel (NSHM) was classified as a category 5 stock, based on the ICES approach to data-limited stocks (DLS). Since then, a progressive reduction of TAC was advised by ICES, from 25500 tonnes in 2013-2014 to 15200 tonnes in 2015-2016.

In 2017 this stock was benchmarked and the North Sea International Bottom Trawl Survey (NS-IBTS) and Channel Ground Fish Survey (CGFS) abundance indices where modeled together. The resulting joint index was considered a proper indication of trend in abundance over time and the NSHM stock was upgraded to category 3. Stock advice for NSHM is biannual. In 2017 the advice for years 2018 and 2019 was provided. The joint abundance survey index indicated a continuation in the increasing trend observed since 2013. Nevertheless, the joint survey index for 2017 has shown a sudden change and steep decline, due to the drop of the CGFS survey index, WGWIDE decided to continue with the current advice of 17517 tonnes for 2019.

The data used as input to the NSHM assessment is as provided to the stock assessor by the stock and survey coordinators. The assessment and forecast appear to have been run in accordance with the stock annex.

## For single stock summary sheet advice:

Short description of the assessment: extremely useful for reference of ACOM.

- Assessment type: update (benchmarked early 2017)
- Assessment: analytical category 3 (survey based method)
- Forecast: not presented
- Assessment model: JAXass model. Separable VPA type model
- Data issues: Data available as described in stock annex. Considerable uncertainties may be present in both survey indices, as well as in catch and bycatch statistics within and between years. Marked decline in the exploitable biomass index, mostly due to the decrease in the CGFS index. Signals of lower recruitment in the English Channel in 2017 compared to 2016. Steps may need to be taken next year for NSHM, if this steep decline continue, to ensure that the stock is kept in a healthy state and fished sustainably.
- Consistency: This years's assessment has been conducted in a manner consistent with last year (benchmark) and stock annex.
- Stock status: F/Fmsy slightly above 1
- Management Plan: No


## General comments

The assessment is well documented and structured. It is quite easy to follow. Applying CPUE from the fishery is not optimal as input data for stock abundance and may involve uncertainties not possible to properly identify and quantify.

## Technical comments

The assessment is done according to decisions taken during benchmark in 2017 and according to the stock annex.

## Conclusions

The updated assessment has been performed correctly. Stock advice for NSHM is biannual. Given the steep decline in the index documented in 2017 compared to 2016 during the second year of an advice, care should be taken when establishing biannual advice for NSHM in the future.

## Audit of Boarfish

Date: 2018.09.13
Auditor: Sólvá Káradóttir Eliasen (input data and assessment)

## For single stock summary sheet advice:

- Assessment type: update
- Assessment: trends - Category 3 stock
- Forecast: not presented
- Assessment model: Bayesian Schaefer state space surplus production model fitted using catch data, 6 delta-lognormal estimated IBTS survey indices, and 1 acoustic survey estimate. Key parameters (r, K, Fmsy, Bmsy and TSB) have been estimated using the exploratory Schaeffer state space surplus production model. The assessment has been run by the WinBUGS14 program.
- Data issues: Input data (i.e. yearly total biomass derived from acoustics, annual total catches and survey data) are available on Sharepoint as described in the Stock Annex. Catch and acoustic biomasses are also available in the WGWIDE report and in stock annex. There are inconsistencies between assessment input in landings/discards/catch data (catch.data.xlsx) and landings/discards/catch data in table 3.1.2.1. However, this does not affect the assessment, since the total catches are correct. This issue will be clarified in next year's report. The survey data are only available in the data folder, and thus it is not possible to double check whether these are consistent.
- Stock status: <Bmsy>=1.63e5 < TSB. $50=2.5 \mathrm{e} 5$ and $<$ Fmsy $>=0.18$. F for 2018 has not yet been estimated, but it was 0.056 in 2017.
- Management Plan:


## General comments

In general, the assessment model is well described.

## Conclusions

The assessment has been performed correctly

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice?
- Yes
- Is the assessment according to the stock annex description?
- Yes
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary?
- Have the data been used as specified in the stock annex?
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
- Is there any major reason to deviate from the standard procedure for this stock?
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice?


## Audit of Red gurnard

Date: 07.09.2018
Auditor: Konstantina Dimitrakopoulou

## General

Survey data available in DATRAS was reported for this stock for the IBTS-Q1 survey in the North Sea, the WCGFS Scottish West Coast Groundfish Survey, the IGFS Irish Groundfish Survey, the French EVHOE-WIBTS-Q4 in the Celtic Sea and Bay of Biscay and CGFS-Q4 in Division 7d. Survey abundance information was provided separately for SP-PORC and SP-NSGFS, the Spanish Porcupine and Northern Spanish groundfish surveys.

The landings data are not species-specific in the fisheries and there are currently no technical measures specifically for managing the fishery. There is need for regular sampling of red gurnard in commercial landings and discarding to provide series of length or age compositions to conduct analytical assessment.

## For single stock summary sheet advice:

- Assessment type: update
- Assessment: not presented
- Assessment model: NA
- Data issues: landings data are not species-specific, lack of biological sampling in commercial landings and discarding
- Consistency: NA
- Stock status: Uncertain
- Management Plan: NA


## General comments

This is a well-documented section.

## Technical comments

None

## Conclusions

The assessment has been performed correctly.

# Audit of Blue whiting (Micromesistius poutassou) in subareas 27.1-9, 12, and 14 (Northeast Atlantic) 

Date: September 6, 2018
Auditor: Anna H. Olafsdottir

## General

Assessment model, recruitment estimates, and forecast model were executed according to stock annex description. The updated assessment gives a valid basis for advice.

## For single stock summary sheet advice:

- Assessment type: update. Benchmark in 2012 (WKPELA 2012) and adaptation of model at an inter benchmark in 2016 (IBPBLW 2016).
- Assessment: age-based analytical assessment that uses catches both in the model and the forecast.
- Forecast: presented.
- Assessment model: SAM with length and age composition of commercial catch and as tuning series an age segregated tuning index from scientific acoustic survey of the spawning stock in March on main spawning grounds on the shelf west of Ireland.
- Data issues: input data for the assessment, as described in the stock annex, are available online at https://www.stockassessment.org/index.php in folder "BW_2018". The mean weight at age from the Faroese catch-at-age data from 2016 were brought up for discussion at this years assessment. Exploratory runs indicated some discrepancy when included in the assessment. It was decided to await for updated age-readings at next year's assessment, before any changes will be made to the input data
- Consistency: assessment results for SSB and F show a decline in SSB and increase in F compared to last year assessment.
- Stock status: $\mathrm{SSB}>\mathrm{B}_{\text {trigger }}$ and $\mathrm{F}_{\mathrm{MSY}}<\mathrm{F}<\mathrm{F}_{\text {pa }}$. Trend in recruitment low for the last two years.
- Management Plan: Long-term management strategy agreed in 2016. The main elements of the plan is catch set at FMSY when SSB forecast $\geq B_{\text {trigger, }}$, reduced F when $\mathrm{B}_{\text {trigger }}>\mathrm{SSB}>\mathrm{B}_{\mathrm{lim}}$, and set $\mathrm{F}=0.05$ when $\mathrm{SSB}<\mathrm{Blim}_{\text {lim }}$. There are $20 \%$ reductions and $25 \%$ increase contain on annual deviation in TAC. Plan is evaluated by ICES and regarded consistent with the precautionary approach.


## General comments

This was a well documented, well ordered, short and to the point section. It was easy to follow and interpret. There were minor discrepancies between subchapters number and numbering of figures and tables.

## Technical comments

- Consider supplying the results of the forecast and annual catch scenarios online with the assessment.
- There are minor errors in text on the advice sheet: 1) in section "Basis of the assessment" length frequency of catch data is listed as input data. Also,
weight-at-age in the catch is missing from the input data list; 2) in section "Indicators" list of surveys used to evaluate recruitment for age-1 and age2 is wrong. According to presentation at WGWIDE and the stock annex, the IESSNS and EVHOV are not used to evaluate recruitment.
- There are minor issues with numbering of figures and tables in the report text. Numbering of figures and tables is not coordinated with subchapter numbering in report text. This applies to all figures and tables from report subchapter 2.4.1 and onward. Labelling of some figures and tables is not in chronological order in the report text. For example, in report text reference to Table 2.3.1.2.5 before Table 2.3.1. This occurs at several occasions of figures and tables in the report. Tables 2.3.1.4, 2.3.2.1.3, 2.4.2.4, and 2.4.2.6 are not referred to in report. Figure 2.3.1.5. is not referred to report.
- Minor discrepancies between report text and data reported in tables: 1) in sub-chapter 2.4.1.1: sampling intensity in report text missing a few areas; 2) in sub-chapter 2.4.2 report text has preliminary catches for Q1 and Q2 in 2018 as 1351802 ton compared to 1330754 in Table 2.3.2.1.
- Minor mistakes in Tables and Figures: 1) in Table 2.8.2.2.1: when F=0, catch in 2019 is listed as 4 ton; 2) Figure 2.9.1 is labelled as displaying the period from 2010 to 2018, however it appears to display assessments from 2013 to 2018.


## Conclusions

The assessment has been performed correctly

## Blue whiting.

Date: 2 September 2018
Auditor: Nikolay Timoshenko

## General

WG suggests that the catches in 2018 should be no more than 1712870 tonnes. The assessment is based on knowledge of the level and structure of the catch in the first half of year. Proportion of the annual catch-at-age taken in the first semester of 2015-2017 was used for raising the preliminary first half year of 2018 catch data. Such predicts have not so far been accompanied by notable deviations and seem acceptable to be applied in the cohort programs. BWSSS provides the basis of fitting which from two youngest age groups are excluding. Comparison with the results of other surveys convinces that as data accumulates, it will be possible to return to this question. In general, the assessment is satisfactorily provided by the input data.

## For single stock summary sheet advice:

The evaluation methodology was described in the previous reports of WGWIDE.

- Assessment type: update/SALY
- Assessment: analytical
- Forecast: presented
- Assessment model: SAM, TISVPA, XSA +1 survey
- Data issues: The data for 2017 presented completely in the annex. Data for 2018 in part were as the results of the assumptions.
- Consistency: The view of the WG was that last years assess should have been accepted.
- Stock status: B is clearly more than Bpa. F<Fpa. R seems to be low last years.


## General comments

Report is well documented, contains relevant explanations and references. Assessment provides a valid basis for advice. The contents of the report correspond to the agenda. The data been used as specified in the stock annex. There is no reason to deviate from the standard procedure for this stock. Reliable recruitment forecast remains to be as the main task.

## Technical comments

The three models applied show similar dynamics in biomass and recruitment. The SSB values are estimated to be increasing in 2011-2018. That growth potential is corresponding by the presence of strength generations of 2013-2015. Later, the growth of biomass has ceased to prevail over its decline in accordance with the conclusion of WG last year. Dynamics of F is also the same in all models until 2017. In 2018, SAM and TISWPA show a decrease in $F$ in the age range $3-7$ while XSA records a slight increase. However, if the range is extended to the ages of $1-8$, XSA also shows a decrease. Such differences are due to the difference in the selection pattern.

## Conclusions

The assessment has been performed correctly. The dynamics of the blue whiting stock has been described by the fishing mortality exceeding $\mathrm{F}_{\text {msy }}$ for a long time. The biomass
remains noticeably higher than the corresponding reference points. That means that the chosen strategy facilitated the retention of the SSB in precautionary boundaries. The detected decline in biomass will require a more careful attitude to the recommendations of the group in respect of following the Fmsy rule.

## Audit of (Stock name)

Date: 02-09-2018

Auditor: Claus Reedtz Sparrevohn

## General

This audit is written for the use during the ADG. There has been no deviation from the stock annex and the assessment are much in line with the previous 2018 assessment.

This assessment only prompted few discussions during the meeting. That the 2019 advice was smaller than the 2018 advice was a bit surprising giving that the IBWSS index was historical high. However, the EG agreed that this could fully be explained by 1) the higher TAC (+325 002 tons) and 2) the small incoming year classes (2015 and 2016). It should be noted by the ADG , that in the advice it is stated that the recruitment in both 2016 and 2017 was low, although in the report it is only stated that the 2017 is low (section 2.1). The size of the 2016 recruitment (2015 yearclass) can be discussed.

As part of the audit the numbers presented in the advice has been check with the numbers that appear in the report. Some small discrepancy:

- Total 2017 catches is 1558061 kt in the rapport and 1555069 in table 12 in the advice sheet.


## For single stock summary sheet advice:

Short description of the assessment: extremely useful for reference of ACOM.

- Assessment type: update
- Assessment: analytical
- Forecast: presented
- Assessment model: SAM
- Data issues: One survey (IBWSS) used. Catch data for 2018 estimated by raising the quarter 1 and 2 catches.
- Consistency: This year assessment is basically in line with last year assessement
- Stock status: Above MSY Btrigger
- Management Plan: Agreed in 2017. In the management plan a stability clause $(-20 \% /+25$ TAC constraint) is set out. The plan is evaluated by ICES assuming that catches will equal the advised management plan TAC. This is not the case for 2018, where the total catch is assumed to equal the sum of national quotas, which is $23.4 \%$ higher, that the advice when applying the Management Plan. Therefore, the EG was in agreement that the - $20 \%$ should be calculated from the latest ICES advice and not the TAC. This lead to a decrease on $-17.6 \%$ and hence the TAC constraint was not considered relevant.


## General comments

The assessment was performed correctly and did not prompt much discussion at the meeting.

## Technical comments

None

## Conclusions

The assessment has been performed correctly

## Checklist for audit process

General aspects

- Has the EG answered those TORs relevant to providing advice?
- Yes
- Is the assessment according to the stock annex description?
- Yes
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary? - Yes
- Have the data been used as specified in the stock annex?
- Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
○ Yes
- Is there any major reason to deviate from the standard procedure for this stock?
- No
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice?
- Yes


## Striped red mullet.

Date: 6 September 2018
Auditor: Anatoly Chetyrkin

## General

For single stock summary sheet advice:

- Assessment type: no assessment due to lack of age structured analytical input data
- Assessment: Not presented
- Forecast: Not presented
- Assessment model: None
- Data issues: General lack of data, both sampling and time-series.
- Consistency: NA
- Stock status: undefined.
- Management Plan: undefined.


## General comments

This is a well documented section, but the lack of information and data omit any usable conclusion and advice on this species.

## Technical comments

The total number in the table 10.1 for 2006,2008,2011,2013,2014 and 2017 ** years does not coincide with the sum of the columns by 1 . This is certainly a rounding problem in the calculation. For 2017* there is an error in total value calculation.

The total number in the table 10.2 for 2007,2010-2014,2016 and 2017 ** years does not coincide with the sum of the columns by 1 . This is certainly a rounding problem in the calculation. For 2008 the total value is calculated incorrectly.

There is no reference to Table 10.2 in the text.

## Conclusions

The assessment has been performed correctly when a few corrections have been made in the tables.

## Audit of Striped red mullet

Date: 07/09/2018
Auditor: Patrícia Gonçalves

## General

Age structured analytical model is not possible due to short time-series of available data.

## For single stock summary sheet advice:

- Assessment type: update/SALY There is no assessment, due to a short timeseries of age data available.
- Assessment: limited data available to evaluate stock trends.
- Assessment model: no assessment.
- Data issues:
- Consistency:
- Stock status: undefined.
- Management Plan: there is no management plan.


## General comments

The section is well structured.

## Technical comments

The 2006 total landings are different in Table 10.1 and Table 10.2
Table 10.2 is not mentioned on the text.
The cited references: Jones, 1972; Russel, 1976; are not included on the references list.

## Assessment type: update Western horse mackerel (hom.27.2a4a5b6a7a-ce-k8) - data audit

Date: 8 September 2018

Auditor: Martin Pastoors

## General

The Western horse mackerel assessment has been carried out using Stock Synthesis 3.30. This audit only focusses on the data that is being used for the assessment.

When auditing the input and output data to this assessment, it was noticed that the tracking of the data throughout the assessment process is quite challenging. Input datafiles are prepared specifically in the format required by Stock Synthesis, however the link between the basic input data and the input file for the assessment needs to be better documented and explained. Ideally, the input data should be available in standard readable formats so that other assessment models than Stock Synthesis could also be deployed.

The assessment itself is consistent with the assessment carried out in 2017, although the retrospective upward revision of biomass and downward revision of fishing mortality has again occurred this year.

## Summary

- Assessment type: update/SALY
- Assessment: analytical
- Forecast: presented
- Assessment model: Stock Synthesis 3.30
- Data issues: The main issue with the data for this assessment is the difficult in tracking the different sources of input data and how they lead to the Stock Synthesis input file. It is recommended to provide a detailed step-by-step documentation how the data is being worked up. In the current situation it is not feasible to completely check derivation of the input data to the stock assessment from the raw data files.
- Consistency: The view of the WG was that the assessment should be accepted. However, there was a major discussion on the applicability of the biomass reference points which were estimated at the benchmark in 2017. Due to the retrospective revisions after the benchmark, the stock size over a period of around 15 years has been estimated to be higher than in the previous assessments. Because the Blim was set as the Bloss of the benchmark assessment in 2017 which also happened to be the last data point in the time series, the applicability of the biomass reference points was seriously questioned. An interbenchmark has been proposed to address this issue.
- Stock status: B is between Blim and Btrigger. F is well below Fmsy.


## General comments

The report is well documented and contains relevant explanations and references in line with the reports of previous years. The assessment has been used as the basis for the advice although concerns have been raised in the WG about the applicability of the biomass reference points or on the question whether the assessment should be used as an absolute or relative indication of development of the stock. Given that this was an
update assessment, in the end the stock annex was followed which resulted in the advice that is in the draft advice document. The data been used as specified in the stock annex although, as mentioned above, the documentation of the input data is difficult to track. Reliable stock indicators remain an important challenge for the assessment, since there is only the egg survey (every three years), a recruitment index and a biomass and length-frequency index from the southern part of the distribution area.

## Technical comments

Only one model (Stock Synthesis) has been applied to this stock as specified in the stock annex. Previously the stock was assessed with the SAD model, but the development of that model has been terminated. Stock Synthesis requires two input files: a control file and a data file. The control file contains the settings to be used in the model and also the values of the assumed variables like natural mortality ( $\mathrm{M}=0.15$ for all ages and years).

The data file contains a specification of the datasources that are being used and the actual data series. Data series that are not used in the model but instead are calculated (e.g. maturity, weight, fecundity) are not included in the data file even though that data may be available in the underlying data sources.

SSB is around the lowest of the time series but recruitment appears to have been a bit higher over the past few years. Fishing mortality is estimate around 0.06 in the most recent year which is substantially lower than the Fmsy. The retrospective revisions of the stock estimates have been a feature of the western horse mackerel assessment for many years already. Unfortunately, the Stock Synthesis model does not seem to have remedied that situation.

## Conclusions

The assessment has been performed according to the specifications in the stock annex. Concerns have been raised about whether the assessment is capable of measuring the absolute level of biomass and fishing mortality of this stock. The biomass is now estimated to be close to MSY $B_{\text {trigger }}$ by the virtue of the retrospective revisions of the assessment relative to the fixed reference points. An interbenchmark has been proposed to address this issue. The interbenchmark could also explore the potential application of a second assessment model as a confirmation of the trends observed in Stock Synthesis.

The documentation and transparency of the input data for the assessment needs to be improved.

## Audit of mac.27.nea

Date: 3rd September 2018
Auditor: Andrew Campbell

## General

The WG accepted the update assessment as a basis for advice for 2019 but is concerned with some aspects of the data and assessment model. The assessment is particularly sensitive to the inclusion of an additional year of RFID tagging data. However, the group could find no compelling argument to exclude this data from this update assessment. A need for improved understanding of model behavior the development of additional model metrics to investigate the weighting given to individual datasets prompted the group to propose ToRs for an interbenchmark exercise.

The fishery independent datasets currently indicate a declining stock, which, combined with high catches assumed for 2018 lead to a predicted SSB below MSY Btrigger ( 2.57 Mt ) in 2018.

There are possible issues with over parameterization of the assessment model with some strong correlations between parameters are noted.

The data used as input to the assessment is as provided to the stock assessor by the stock and survey coordinators. The mechanism for the delivery of this data requires formalization for auditing and quality checking purposes. The assessment and forecast appear to have been run in accordance with the stock annex.

## For single stock summary sheet advice:

- Assessment type: update (benchmarked early 2017)
- Assessment: analytical, category 1
- Forecast: presented
- Assessment model: SAM, modified to utilise tag/recapture dataset. FLR forecast.
- Data issues:

No recruitment index is available for 2016 and 2017. This necessitates a departure from the procedure outlined in the stock annex for estimating the recruitment estimates required for the short term forecast, whereby the terminal assessment year recruitment estimate is replaced.

There were minor updates to the historical RFID tagging dataset although changes to recapture rates were generally $<1 \%$.

Some other issues were clarified by the stock assessor during the audit and did not necessitate any changes to the assessment.

- Consistency: This year's assessment has been conducted in a manner consistent with last year and the stock annex. Outputs indicate revisions to absolute SSB and F over the last 10 years of the order of $10 \%$.
- Stock status: SSB is forecast to fall below MSY Btrigger (2.57Mt) in 2018 if the intermediate year catch (approx. twice the advice) is realised. Intermediate year catch assumptions in previous years have proved accurate. Maintain-
ing the current catches (approx. 1Mt) into 2020 would result in the SSB falling below Blim in 2020. Fishing mortality has been increasing since 2011 and exceeds $\mathrm{F}_{\mathrm{pa}}(0.35)$ in 2017.
- Management Plan: ICES advised on a proposed management plan from EU, Norway and Faroe Islands in September 2017 (and also revised the fishing mortality reference points). A suite of target fishing mortality and biomass trigger points were evaluated for a hockey-stick type HCR. The requesting parties agreed on a LTMP with a target F of 0.21 and a trigger point of 2.57 Mt (coinciding with the MSY reference points). However, since not all fishing parties are in agreement, ICES advice is based on the MSY approach.


## General comments

The draft report text that was available at the time of this audit is well structured and clear. The assessment code is relatively clear and concise. However, a number of assessment model parameter settings are not explicitly detailed in the code (likely because they assume default values). An explicit line of code/comment would aid auditing.

Inclusion of code for tabulating the assessment output would be beneficial for auditing purposes.

The STF code was slightly more complex. IBTS index time series is hard-coded in script and ideally should be input from the same file as used for the assessment script.

## Technical comments

The IESSNS catchability parameter couplings (-1/-1/-1/3/4/4/4/4/4/4/4/4/-1) do not match those in the stock annex ( $-1 /-1 /-1 / 2 / 3 / 4 / 5 / 6 / 7 / 8 / 9 / 9 /-1$ ) or the draft report but they are consistent with the 2017 assessment.

The stock annex includes a clear explanation with regard to the calculation of the recruitment estimate in the terminal assessment year. However, for both this and last year it has not been possible to follow the procedure and an alternative has been used. This should be described in the stock annex along with an exact specification of the years to be used when calculating the geometric mean for the recruitments in the period of the short term forecast.

## Conclusions

The assessment has been performed correctly.

## Checklist for audit process

General aspects

- Has the EG answered those TORs relevant to providing advice?
- Is the assessment according to the stock annex description?
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary?
- Have the data been used as specified in the stock annex?
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
- Is there any major reason to deviate from the standard procedure for this stock?
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice?
1 ) /SALY There is no assessment, due to a short time-series of age data available.

2 ) Assessment: limited data available to evaluate stock trends.
3 ) Assessment model: no assessment.
4) Data issues:
5) Consistency:

6 ) Stock status: undefined.
7 ) Management Plan: there is no management plan.

## General comments

The section is well structured.

## Technical comments

The 2006 total landings are different in Table 10.1 and Table 10.2.
Table 10.2 is not mentioned on the text.
The cited references: Jones, 1972; Russel, 1976; are not included on the references list.

## Audit of Mackerel (WGWIDE 2018)

Date: 7. September 2018
Auditor: Jan Arge Jacobsen

## General

The stock assessment for NEA mackerel in 2016 has been done according to the stock annex. The stock is estimated to be below MSY Btrigger in 2018, for the first time since 2007, and the advice is thus based on the MSY approach: FMSY * SSB(2019) /MSY B trigger

## For single stock summary sheet advice:

Short description of the assessment:

- Assessment type: update
- Assessment: analytical
- Forecast: presented
- Assessment model: State-space model (SAM) fitted to catch-at-age data for ages 0 to 12 (+ group) (1980-2017, strongly down-weighted for 1980-1999) and three surveys: 1) Mackerel Egg survey (triennial, 1992-2016); 2) Recruitment index from IBTS Q1 and Q4 surveys (1998-2015); and 3) abundance estimates, ages 6 to 11, from IESSNS survey (2007 and 2010-2018). The model also incorporates tagging-recapture data from the Norwegian tagging program (1980-2005) and the new RFID tagging series (2011 and onwards).
- Data issues: New survey input data for the assessment, as described in the stock annex, were available for the IESSNS, tagging-recapture data from the Norwegian tagging program and egg survey series. However, no data was available for the IBTS recruitment index from the North Sea for the second year in a row. With the addition of 2017 catch-at-age, weights-at-age in the catch and in the stock, maturity ogive and proportions of natural and fishing mortality occurring before spawning.
- Consistency: Last year's assessment was accepted
- Stock status: $\mathrm{B}<\mathrm{B}_{\mathrm{mSY}}$ Btrigger and $\mathrm{B}_{\mathrm{lim}}<\mathrm{B}^{2}<\mathrm{B}_{\text {pa }}, \mathrm{F}>\mathrm{F}_{\text {mSY }}$ and $\mathrm{F}_{\mathrm{pa}}<\mathrm{F}<\mathrm{F}_{\text {lim, }} \mathrm{R}$ has bee high since early 2000s but the 2015 and 2016 year-classes are estimated to be below average.
- Management Plan: There is no agreement on an overarching management plan for mackerel. ICES have based their advice on the MSY approach. However EU, NO and FO agreed in 2014 on an ad hoc management plan for the years 2015-2018. The ad hoc Management Plan was evaluated by ICES in 2017 after the benchmark, and was adjusted accordingly for the updated reference points by the three parties for the 2018 advice (refer to Table 8.2.4.1 in the WG report).


## General comments

The sections were well ordered, however not all were finished by the time of the audit. This did not affect the main conclusions. Analyses were well described and the results presented clearly. The conclusions regarding advice are appropriate, given the divergent survey trends, increased reliance on catch data and associated change in perception of stock status. The short time series of some of the survey caused instability in the model, as the "leave out" runs clearly demonstrated. The perception of the stock
changed proportionally much by on removal of single input series. The model might be over parameterized.

## Technical comments

Due to missing IBTS recruitment index the setting for the SAM run were not entirely in accordance with stock annex (benchmark 2017). The outdated time series from WGWIDE 2016 was used in the assessment and it was therefore conducted without an index value for the 2016 and 2017 year classes (see Sec. 8.6.2 for details).

There might be some issues with the use of the tagging data (RFID) in the SAM model. In Section 8.6.4 a discussion of possible effects/biases were discussed that might have large influences on the model runs, e.g. if the recapture rates in different areas/seasons vary due according to incomplete mixing of tagged fish, or due to mortality happening between seasons (for instance between quarter 1 and quarter 4 catches). These potential biases are not taken into account in SAM assessment model today. Also the high tagging mortality for the RFID tags that the model estimates were considered problematic in the assessment. Alternative use of tag data in the assessment was discussed in this section. An ICES workshop on tagging data was suggested.

The realisation of the process error in the model was also inspected (Section 8.7.5). While process error is assumed to be independent and identically distributed, there is clear evidence of correlations in the realisation of the process error in the mackerel assessment, which appears to be correlated both across age-classes and temporarily. The temporal autocorrelation can also be visualised if the process error is expressed in term of biomass (deviations in abundances-at-age multiplied by weight at age and summed over all age classes, Figure 8.7.5.4 in the WG report). For the years since 2010 the cumulated process error remains positive, with the magnitude reaching a third of the volume of the catches for 2009. The reason for this misbehaviour of the model could not be identified.

## Conclusions

The update assessment has been performed correctly and gives a valid basis for advice. The WG recommends an interbenchmarch as soon as possible (2019) to deal with data/model issues.

## Annex 6: List of Working Documents

i) Cruise report on the International ecosystem survey in Nordic Seas (IESNS) in May - June 2018. Authors: ICES 2018 with contribution from Maxim Rybakov, Tatyana Sergeeva, Anna Gordeeva, Valantine Anthonypillai, Are Salthaug, Erling Kåre Stenevik, Kjell Arne Mork, Cecilie Thorsen Broms, Øystein Skagseth, Karl-Johan Stæhr, Benoît Bergès, Mathias Kloppmann, Sven Kupschus, Guðmundur J. Óskarsson, Anna Heiða Ólafsdóttir, Hildur Pétursdóttir, Eydna í Homrum, Ebba Mortensen, Sólva Eliassen, Poul Vestergaard, Leon Smith.
ii) Herring distribution in the Norwegian Sea in May. Authors: S. K. Eliasen, E. í Homrum, and J. A. Jacobsen.
iii ) Pelagic ecosystem acoustic-trawl survey PELACUS 0318: mackerel, horse mackerel, blue whiting and boar fish abundance estimates. Authors: Pablo Carrera, Paz Díaz, Rosario Domínguez, Gonzalo González-Bueno, Isabel Riveiro
iv) Pelagic ecosystem acoustic-trawl survey PELACUS-IBWSS 0318: blue whiting and müeller's pearlside fish abundance estimates in Porcupine Seabight. Authors: Pablo Carrera, Paz Díaz, Rosario Domínguez, Gonzalo González-Bueno, Isabel Riveiro
v ) Cruise report from the International Ecosystem Summer Survey in the Nordic Seas (IESSNS) 30th of June - 6th of August 2018. Authors: Anna Heiða Ólafsdóttir, Sigurður Pór Jónsson, James Kennedy, Jan Arge Jacobsen, Ebba Mortensen, Leon Smith, Teunis Jansen, Søren Post, Lars Heilmann, Kjell Rong Utne, Leif Nøttestad, Valantine Anthonypillai, Are Salthaug, Åge Høines, Geir Odd Johansen, and Kai Weiland.
vi ) Comments on incongruous formulations in the SAM (state-space assessment model) model and consequences for fish stock assessment. Authors: M. Aldrin, S. Aanes, S. Subbey (Reference: M, A., Fisheries Research (2018), https://doi.org/10.1016/j.fishres.2018.08.001).
vii ) Reference fleets identification by LPUE data filtering applied to the striped red mullet (Mullus surmulletus) in the Bay of Biscay. Authors: Nathalie Caill-Milly, Muriel Lissardy, Noëlle Bru, Marie-Adèle Dutertre, Cassandre Saguet.
viii ) Red gurnard in DCF/NP samplings for ICES Division 27.9a. Authors: Diana Feijó and Alberto Rocha
ix ) Gurnards: species landings' composition in ICES Division 27.9a. Authors: Alberto Rocha, Diana Feijó and Patrícia Gonçalves.
x ) NEA mackerel, alternative assessment. Authors: Höskuldur Björnsson
xi ) Direct assessment of small pelagic fish by the PELGAS18 acoustic survey. Authors:Duhamel, E., Doray, M., Huret, M., Sanchez, F., Marie-Lepoittevin, T., Peltier, H., and Autthier, M.
xii ) Norwegian Spring Spawning Herring stock assessment by means of TISVPA. Authors: Dimitry Vasilyev
xiii ) Survey report for MS Eros, MS Kings Bay MS Vendla 13.-25.02.2018 in relation to Distribution and abundance of Norwegian spring-spawning herring during the spawning season in 2018. Authors: Aril Slotte, Are

Salthaug, Åge Høines, Erling Kåre Stenevik, Sindre Vatnehol and Egil Ona.
xiv ) Blue Whiting stock assessment by means of TISVPA. Authors: Dimitry Vasilyev.
xv ) Evaluation of potential rebuilding strategies for the Western Horse Mackerel (Trachurus trachurus) fishery. Authors: S.P. Cox, A.J. Benson, B. Doherty, and S. Johnson.
xvi ) Developing a standardized Horse mackerel CPUE index. Authors: Esther Beukhof and Martin Pastoors.
xvii ) PFA self-sampling report for WGWIDE (2015-2018). Authors: Martin Pastoors.
xviii )2018 mackerel egg exploratory survey. Authors: Finlay Burns, Brendan O'Hea, Bjorn Gunnarsson
xix ) Nøttestad, L. Utne, K.R., Sandvik, A., Skålevik, A., Slotte, A. and Huse, G. 2018. Historical distribution of juvenile mackerel northwards along the Norwegian coast and offshore following the 2016 mackerel spawning. Working Document to ICES Working Group on Widely Distributed Stocks (WGWIDE), Havstovan, Tórshavn, Faroe Islands, 28. August - 3. September 2018, 25 pp.
xx ) Authors: Leif Nøttestad, Kjell Rong Utne, Anne Sandvik, Åsmund Skålevik, Aril Slotte, Geir Huse


[^0]:    * Discards data from UK(Scotland) were provided by year, due to sampling intensity.

[^1]:    * Data from UK(England + Wales) not included (2004-2007).
    ** Data from UK(England + Wales) and Sweden not included (2008-2011).

[^2]:    * landings only.
    ** those values are assumed, since data of catches inside/outside NEAFC was not available.

[^3]:    * (final-preliminary)/final ${ }^{*} 100$

[^4]:    * $95 \%$ confidence interval
    ** compared to sum of national quotas in 2017, not advice for 2017
    *** difference in fourth decimal compared to $F$ status quo

