# ICES HAWG Report 2007 

ICES Advisory Committee on Fishery Management ICES CM 2007/ACFM:11

# Report of the Herring Assessment Working Group South of $62^{\circ} \mathrm{N}$ (HAWG) 

13-22 MARCH 2007
ICES HEADQUARTERS

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## Executive Summary

The ICES herring assessment working group (HAWG) met for 10 days in March 2007 to assess the state of 7 herring stocks and 3 sprat stocks. New data were only available for 6 herring stocks and 2 sprat stocks. HAWG carried out a benchmark assessment on Celtic Sea herring. The following issues were explored:

- catch data through catch curves
- simpler models using CSA
- standard catch at age assessment models (ICA and XSA)
- time inconsistencies, outliers and time trends in indices.
- retrospective performance of the different models.
- Recruitment and productivity changes

The exploration showed that there is uncertainty in SSB, F and recruitment for last 3 years in Celtic Sea herring. However information from the catch shows an increasing trend in the mortality of the fish and a contraction in age structure of the stock. Exploration with simpler models showed a decline in biomass over the whole last 10 years. A Bayesian analysis suggests that the selection of the fishery has changed over the last 10 years, and supported the perception that the current status of SSB is uncertain, but probably at a low level. Analysis of recruitment patterns suggested that no major regime shift has taken place in Celtic Sea herring productivity in the last 40 years.

The recent trends in North Sea autumn spawning herring show that after a peak in spawning biomass (SSB) of 1.8 million tonnes in 2004, the SSB in 2006 was 1.2 million tonnes. The current fishing mortality $\left(\mathrm{F}_{2-6}\right)$ is 0.35 and is well above the target F prescribed by the management agreement. It is likely that the stock will decline further in the next few years to close to $\mathrm{B}_{\text {lim }}$ by 2009. The decline in SSB is due to serial poor recruitment since 2001 and a failure to fish adult herring at target $\mathrm{F}(0.25$, as described in the management agreement) in the last few years. The estimate for the most recent recruiting year class is the lowest since 1979 and the low recruitment is caused during the larvae phase of North Sea herring.

All herring stocks assessed by the working group appear to have average or below average recruitment in the last few years. The fishing mortality on herring in IIIa is now considered too high, especially in light of MSY targets. This is also the case for herring for the west of Scotland (VIaN). There is no sign of stock recovery in VIaS herring. Conflicts in the data, made it impossible to assess Irish Sea herring, although data exploration suggest that the age profile of the stock has contracted and the SSB is stable at a low level. It is likely that the abundance of North Sea sprat is now less than in the last two previous years.

HAWG answered one special request from the EU on the findings of the WESTHER project, particularly with reference to the proposed management plan for herring to the west of Scotland (VIaN). See section 1.3 for the full answer.

HAWG also commented on the quality and availability of data, the problems with estimating the amounts of discarded fish, the use of the new data system INTERCATCH, the relevance of ecosystem changes to the stocks considered by the group and recent meetings and reports of relevance to HAWG. An analysis of the surplus and net production of 5 herring stocks also found that fisheries-independent shifts in productivity had occurred since the 1960s in North Sea, west of Scotland and Irish Sea herring.

HAWG was concerned about the apparent increase in misreporting of catches in recent years and the growing relaxation of regulations designed to restrict the ability to misreport or catch herring as bycatch.

### 1.1 Participants

| Steven Beggs | UK/Northern Ireland |
| :--- | :--- |
| Hans Bogaards | The Netherlands |
| Massimiliano Cardinale | Sweden |
| Maurice Clarke | Ireland |
| Mikael van Deurs | Denmark |
| Mark Dickey-Collas (Chair) | The Netherlands |
| Afra Egan | Ireland |
| Tomas Gröhsler | Germany |
| Joachim Gröger | Germany |
| Emma Hatfield | UK/Scotland |
| Henrik Mosegaard | Denmark |
| Peter Munk | Denmark |
| Mark Payne | Denmark |
| Beatriz Roel | UK/England \& Wales |
| Marine Pomarede | UK |
| Norbert Rohlf | Germany |
| John Simmonds | UK/Scotland |
| Jorn Schmidt | Germany |
| Dankert Skagen | Norway |
| Else Torstensen | Norway |
| Christopher Zimmermann | Germany |
| Yves Verin | France |

Contact details for each participant are given in Annex 1.

### 1.2 Terms of Reference

## 2006/2/ACFM04

The Herring Assessment Working Group for the Area South of $62^{\circ} \mathrm{N}$ (HAWG) (Chair: Mark Dickey-Collas, The Netherlands) will meet in Copenhagen, Denmark from 13 - 22 March 2007 to:
a) assess the status of and provide management options (by fleet where possible) for 2008 for:

- the North Sea autumn-spawning herring stock in Division IIIa, Subarea IV, and Division VIId (separately, if possible, for Divisions IVc and VIId). Forecasts should be provided by fleet if possible and taking into account the management plan agreed between the EU and Norway;
- the herring stocks in Division VIa and Sub-area VII;
- the stock of spring-spawning herring in Division IIIa and Subdivisions 22-24 (Western Baltic); Management options for Division IIIa shall be given by fleets taking into account that North Sea herring and Western Baltic herring are taken together in this Division;
b) assess the status of the sprat stocks in Subarea IV and Divisions IIIa and VIId,e;
c) consider implications of SGRECVAP for the assessment and outlook of North Sea herring stock, as well as for PA reference points;
d) for the stocks mentioned in a) and b) perform the tasks described in C.Res. 2006/2/ACFM01.

HAWG will report by 2 April to the attention of ACFM.

### 1.3 Working Group's response to ad hoc requests

### 1.3.1 Request by European Commission (22/02/2007 D02012) on WESTHER and management of VlaN herring.

ICES received one special request from the European Commission to be considered by HAWG 2007.

## Background

The WESTHER project proposed, tested and reported on one null hypothesis, and three alternative hypotheses, on the structure of herring populations to the west of the British Isles. Their report presents detailed reasoning and conclusions for each hypothesis. To provide the background required for our response, we present here the hypotheses and the WESTHER conclusions. For those who are interested in the supporting arguments we would refer you to the WESTHER report.

The Null hypothesis is that there is only one herring population to the west of the British Isles, with no detectable differences between any of the geographically and temporally separated spawning components. Examination of the null hypothesis involved the consideration of three alternative hypotheses relating to the spawning components, juveniles and feeding aggregations and lead to the following conclusions:

Alternative hypothesis 1: the different spawning aggregations sampled are discrete at spawning time and are, therefore, separate components.

WESTHER found that classification success of spawners was generally high suggesting that there is strong evidence to reject the null hypothesis and accept alternative hypothesis 1 because the different spawning aggregations sampled are discrete at spawning time.

Alternative hypothesis 2: there is clear distinction of juveniles sampled on different nursery grounds.

WESTHER found that parasites and otolith microchemistry act as tags for the juvenile stages of herring. There was a clear distinction between many of the different juvenile samples. There was also strong evidence that juveniles from separate spawning areas mix in some of the nursery areas sampled. WESTHER could distinguish the origin of juveniles even in mixtures, and thus accepted the alternative hypothesis 2.

Alternative hypothesis 3: fish from each spawning aggregation remain discrete on their feeding grounds.

WESTHER rejected alternative hypothesis 3 because there was evidence of mixing of adults from separate spawning components, especially in VIa North. The evidence also suggested that the Celtic Sea and VIIj adults do not mix as much as the more northerly herring. The science, therefore, suggests links between the areas, with fish spawning in different areas mixing, to varying extents, on feeding grounds. However, it was difficult to assess the level of mixing of non-spawning adults.

## The HAWG supports these results and conclusions.

WESTHER briefly examined some of the assessment and management issues that derive from these results and presented the following conclusions:

1. Assess the herring to the west of the British Isles as two stocks - Malin Shelf (including the current ICES stocks VIa North, VIaS and VIIb, c, Clyde and Irish Sea (VIIaN)) and Celtic Sea (the current Celtic Sea and VIIj stock). In the area studied in WESTHER we can hypothesise that there are two stocks within which data can be pooled for assessment. However, the boundary at the northern edge is unclear and there is no evidence presented in the report which separates autumn spawners in the north of Scotland west of $4^{\circ} \mathrm{W}$ from autumn spawning fish east of $4^{\circ} \mathrm{W}$ (the North Sea stock).
2. Survey effort should be increased or diverted to a combined survey on non-spawner distributions mixing on the Malin Shelf.
3. The current monitoring of the spawning components should be maintained, but not to the detriment of a wider scale Malin Shelf survey. Spawning ground surveys might provide data on the dynamics of individual stock components, which are thought to be useful for the development of a fleet-based advice

However,
4. Management plans should be fleet/area based, aiming at preventing the local depletion of any population unit in the area, and should make adaptive changes if current fishing practices change, specifically the introduction of a new $1^{\text {st }}$ or $2^{\text {nd }}$ quarter fishery in the southern part of VIa North and/or northern part of VIaS and VIIb,c.
5. Management plans should recognise the importance of the populations in the north of area VIa as a potential source of herring to spawning grounds to the south.
6. Management plans should recognise that there are potentially two separate stocks on the west coast of the British Isles, these constitute a population in the Celtic Sea and VIIj and a metapopulation centred on area VIa."

HAWG recognizes the need to provide sound management advice for these areas, and in particular the importance of ensuring as far as possible that there is no depletion of local components. However, HAWG noted that WESTHER was not funded to evaluate the extent of mixing in the fisheries or to evaluate alternate management strategies for the area. Currently it is unclear what management regime would provide the most cost effective method for successful management and what data would be needed to support this management.

HAWG considers that it is necessary to move towards an integrated management plan for this area through a series of iterations involving the following steps :-

- Examination of alternative management strategies based on their ability to deliver protection to local populations and provide cost effective information applicable for management of the two stock units of herring to the west of the British Isles.
- Replacement of existing or development of new cost effective assessment and data collection schemes which will be required to support this management.
- Movement to coordinated management for the region.

In this context HAWG proposes a study group with ToR given below.

## Response to Commission

HAWG was requested to "examine the WESTHER report and its recommendations to provide information on necessary changes to ICES long-term management advice concerning the herring stock to the West of Scotland (herring in VIa(N))".

HAWG response: ICES considers that in the absence of any evaluated and coordinated management strategy for the herring to the west of the British Isles, the current separation of management units (VIa(N), VIa(S), Irish Sea and Celtic Sea) affords the best possible protection for local spawning stocks. However it does not afford protection to the fish of one stock distributed in another management area at feeding time.

Provided both the spawning fisheries (VIa(S), Irish Sea and Celtic Sea) and the fishery in the mixing area (predominantly $\mathrm{VIa}(\mathrm{N})$ ) are maintained at an F that would be sustainable for each component, this should afford protection for these units, in the short term. ICES considers that further work is required on examining the issues surrounding surveys, assessment and management of each of the current three management units to the north of the area. This can be initiated, partly through a new study group or study contract. It will be a number of years before ICES can provide a fully operational integrated strategy for these units. In this context ICES recommends that the previously endorsed plans for VIa(N) should be continued, until or unless some alternative strategy is found to be more useful.

## TOR for study group: SGHERWAY

1) Consider the results of WESTHER in relation to VIaN, VIaS and VIIaN stocks.

2 ) Comprehensively evaluate the utility of a synoptic acoustic survey in the summer for the Hebrides, Malin and Irish shelves, in conjunction with PGHERS surveys of VIaN and the North Sea.
3 ) Investigate a alternative assessment methods of the three stocks that take into account WESTHER findings.Investigate their utility for advisory purposes.

4 ) Evaluate, through simulation alternative management strategies for the metapopulation of VIaN, VIaS and VIIaN.
5 ) Comment on what means is best to maintain each spawning component in a healthy state, whilst managing the fish of that component when they are in a neighbouring area.

### 1.4 Reviews of groups or work important for the WG

HAWG was briefed throughout the meeting about other groups and projects that were of relevance to their work. Some of these briefings and/or groups are described below.

### 1.4.1 The Annual Meeting of Assessment Working Group Chairs [AMAWGC]

The working group was addressed by the chair of ACFM (Martin Pastoors). Both he and Mark Dickey-Collas informed the group about the AMAWGC meeting in 2007. They described the construction of a roadmap for the working group for the next 3 years. The HAWG road map can be summarised as:

- 2007- Benchmark assessment Celtic Sea herring, evaluation of Irish management agreements
- 2008- Benchmark assessment of herring in IIIa (western Baltic spring spawning herring), comprehensive descriptions of the fleets
- 2009- Benchmark North Sea sprat

The new developments in mixed fisheries, INTERCATCH, the proposed new ICES advisory structure, evaluation of management strategies and ecosystem descriptions were also discussed at AMAWGC, and were taken into account when the HAWG 2007 report was put together.

### 1.4.2 The Planning Group for Herring Surveys [PGHERS]

The Planning Group for Herring Surveys [PGHERS] (Chair: Norbert Rohlf, Germany) met at the Danish Institute for Fisheries Research, Charlottenlund, Denmark, from 22-26 January 2007 to:
a. combine the 2006 survey data to provide indices of abundance for the population within the area, by means of the FishFrameAcoustics database;
b. coordinate the timing, area and effort allocation and methodologies for acoustic and larvae surveys for herring and sprat in the North Sea, around Ireland, Division VIa and IIIa and the Western Baltic in 2007;
c. intensively test the in-year developments of the FishFrame Acoustics database, specifically verify the ability of the new system to calculate global survey estimates from raw acoustic and trawl data using 2005 and - if possible - 2006 survey data;
d. further harmonise the maturity readings of North Sea and Western Baltic herring conducted by different labs, specifically the definition of mature and immature fish;
e. report on the possible bias introduced by a change in gear in the Dutch herring larvae survey.

Review of larvae surveys in 2006/2007: In total seven units and time periods out of ten were covered in the North Sea. The herring larvae sampling period was finished just prior to the PGHERS meeting, thus sample examination and larvae measurements have not yet been completed. The information necessary for the larvae abundance index calculation will be ready for and presented at the Herring Assessment Working Group (HAWG) meeting in March 2007.

Coordination of larvae surveys for 2007/2008: In the 2007/2008 period, the Netherlands and Germany will undertake seven larvae surveys in the North Sea. Outside the larval sampling programme some additional stations shall be sampled in the area of the Doggerbank to test anecdotal information of a recolonisation of the area and to obtain information about ongoing spawning activity. The Baltic Sea Fisheries Institute will continue with the larvae survey in the Greifswalder Bodden area in 2007, but the survey design will be altered and the N30 time series has to be recalculated to be assessable for the next benchmark assessment of the WBSS.

North Sea acoustic surveys in 2006: Six acoustic surveys were carried out during late June and July 2006 covering the North Sea and west of Scotland. The total combined estimate of North Sea spawning stock biomass (SSB) is 2.1 million $t$. This estimate is comparable to the 1.9 million t SSB in 2005 and the 2004 estimate of 2.6 million t . The stock is dominated by the 2000 year class. Growth of the 2000 year class seems still to be slower than average, individuals of this year class having almost the same size and weights than the one year younger fish of the 2001 year class. The West of Scotland estimate of SSB is 472,000 tonnes ( 190,000 in 2005). This is a substantial increase compared to last years estimate, and the SSB has more than doubled. The SSB is in the same order of magnitude that it had during the last ten years. The surveys are reported individually in Annex 2A-2F.

Western Baltic acoustic surveys in 2006: A joint German-Danish acoustic survey was carried out with RV "Solea" from 05 to 24 October in the Western Baltic. The estimate of Western Baltic spring spawning herring is $214,000 \mathrm{t}$ (compared to 198,000 in 2005). The sprat year class 2006 was overall exceptional strong. In the Kattegat and the northern part of Subdivision 22 anchovy was observed in larger quantities. A full survey report is given in Annex 3.

Manuals for acoustic and herring larvae surveys: Several updates and improvements of the manual for herring acoustic surveys in ICES Divisions III, IV, and VIA have been suggested. The bibliography has been updated accordingly and the list of gears used by the different nations has been updated. The suggested changes are both very relevant and highly needed.

However, the suggested text will need some reviewing which will be done by correspondence in cooperation between Germany, Denmark and Scotland. The revised text will be sent to all members of PGHERS before the next meeting by the Chairman. No modifications were made in the manual for the International Herring Larvae Surveys south of $62^{\circ}$ north.

Status and future of the FishFrame database: All countries have uploaded survey data from 2006 for herring and sprat. The stage 3 dataset for 2003, 2004 and 2005 is completed as well. Testing of the data browser, reports, upload, data checking and interpolation for both in stage 1 and 3 was done during the meeting. Two bugs were found, corrected and testing was redone. FishFrame performed satisfactory and was therefore used to combine the national data into the integrated survey result. FishFrame will be used again for the 2007 survey period

Sprat: Sprat data were available from RV Walther Herwig III, RV Tridens and RV Dana. The total sprat biomass was estimated as $452,000 \mathrm{t}$ in the North Sea (down from 563,000 t in 2005). The biomass is dominated by mature sprat ( $98 \%$ ). The total sprat in the Kattegat was estimated as $33,600 \mathrm{t}$, including $63 \%$ immature sprat (down from $59,800 \mathrm{t}$ in 2005). There is no clear indication that the southern distribution has been reached. However, the highest concentration of sprat was observed off the coast of Scarborough, on the east coast of England.

Coordination of acoustic surveys in 2007: Six acoustic surveys will be carried out in the North Sea and west of Scotland in 2007 between 21 June and 25 July. Participants are referred to Figure 4.3.1.1 for indications of survey boundaries. "Tridens" and "Walther Herwig" will cover the area between $52^{\circ}$ and $57^{\circ}$ together with interlaced transects. A survey of the western Baltic and southern part of Kattegat will be carried out by a German research vessel in October.

Investigation of bias introduced by change in gear in the larvae surveys: In 2004, the Netherlands changed from a Gulf III plankton torpedo to a Gulf VII. However, nothing was known about differences in catchability between these two devices. To investigate the possible bias introduced by the change of gear, real-time fishing comparison trials were conducted in 2006, deploying both samplers in a single frame. Volume filtered by the Gulf VII was found to be significantly higher than in the Gulf III, but catchability was less. However, due to technical problems, no accurate calibration of flow meters could be performed and therefore numbers of larvae caught can not be converted by the volume filtered. During the 2007 mackerel and horse mackerel egg survey, ichthyoplankton hauls with both samplers are planned to compare the catchability of mackerel eggs.

Recent studies on herring fat content and the accuracy of maturity staging: Deborah Davidson from the Aberdeen University gave a presentation of her ongoing PhD study dealing with modern methods of measurements on herring fat content. Data obtained from the 2006 herring acoustic surveys indicate that, as herring length and weight increases, so too does fat content. When comparing fat and weight, fish of a heavier weight did not tend to have very low fat contents. A FATMAP (a visual representation of spatial variation in fat content of North Sea herring) was constructed for the immature herring in the Scotia, Solea and Tridens data. Initial analysis showed that there was a strong linear relationship between length and fat content of immature fish. Plotting the raw fat data showed a clear southeast to northwest trend in fat content.

In a second talk Lindsay McPherson from Aberdeen University presented her findings concerning the accuracy of macroscopic staging of North Sea herring. As the macroscopic maturity scale is based on a histological scale, histology is the most accurate means of assessing maturity stage in fish. New, unambiguous histology keys were formed in order to calculate the accuracy of macroscopic staging. Macroscopic staging on FRV Scotia in 2006 was $78.6 \%$ accurate for females and $83 \%$ accurate for males. Much of the error in female
staging was due to maturing repeat spawners (stage 3 ) being assessed as recovering (stage 8 ). While marked inaccuracies were found they are unlikely to impact upon the SSB as the number of fish assessed as immature (1-2) or mature (stages 3-8) were correct.

## Recommendations:

PGHERS recommended HAWG to comment what proportion of the total survey effort should be directed to the different survey indices (adult herring / sprat / young herring). HAWG stated that it does no recommend any changes in survey effort.

HAWG does not recommend PGHERS to calculate the proportion of skipped spawners (see below).

HAWG recommends PGHERS considering a change in the name for the herring acoustic surveys. Not only herring, but all pelagic fish, especially clupeids like herring, sprat, anchovies and sardines, are monitored in the surveys. The latter ones become more and more abundant in the North Sea. This should be reflected in the labelling of the surveys.

## HAWG response to PGHERS request on Skipped spawners.

A relatively high percentage of herring in the maturity stage 8 (Recovering) was seen in the biological samples from North Sea in the 2006 summer acoustic survey (ICES 2007/LRC:01). PGHERS raised the issue of "Skipped spawners" in the North Sea autumn spawners and has requested HAWG on their needs for data on the proportion of skipped spawners in an assessment context: "HAWG should comment on their needs to calculate a proportion of skipped spawners (Stage 8) for the assessment".

Skipped spawning is considered an important phenomenon in herring (Norwegian Spring spanners) and cod (http://www.ices.dk/iceswork/asc/2004/Theme\ Session.pdf) affecting particularly second time spawners. However, it is unclear how to properly classify skipped spawners. Knowledge of effective fecundity and its annual variations are of great importance in understanding the reproductive potential in a stock. In the assessment of North Sea autumn spawners maturity information (immature and mature) is used in the estimation of the Spawning Stock Biomass.

Differentiating between recruit and repeat spawners is often very difficult in macroscopic determination of the gonads, which is the method currently in use for the North Sea autumn Spawners. In the Norwegian spring spawning herring, there are indication that the spawners are recovering for about three months before they are back in the maturation cycle. It is thus a question if the high proportion of skipped spawners in June-July is related to autumn spawners spawned the previous year. There is little information on mixture of autumn and spring spawners in the different areas of the North Sea.

To include the "skipped spawners" in the current assessment context seems premature at present as available knowledge will most probably not improve the variability in the estimates of SSB. However, the WG recommends that the phenomenon be further studied to improve the methods for maturity determination and the understanding of the relation between recruitment and spawning biomass.

### 1.4.3 Study Group on Recruitment Variability in North Sea Planktivorous Fish [SGRECVAP].

SGRECVAP is due to meet in May 2007 in Plymouth, UK. It will consider the possible causes of the poor herring recruitment in the North Sea, in light of its previous report from 2006, which was discussed in last year's herring working group report. The findings of SGRECVAP have impacted on the choice of recruitment scenarios used for North Sea herring short and medium term projections. As SGRECVAP is yet to meet, HAWG cannot fulfil TOR c.

### 1.4.4 Workshop on Testing the Entrainment Hypothesis [WKTEST]

This workshop will take place in June 2007 in Nantes, France. It will document diagnostic case studies of pelagic fish for the evidence of the entrainment hypothesis and look for understanding of the mechanisms by which life cycles patterns are maintained or changed. It is hoped to used the finding to improve understanding for spatial fisheries management and recovery plans.

### 1.4.5 WESTHER [EU project]

WESTHER: A multidisciplinary approach to the identification of herring (Clupea harengus L.) stock components west of the British Isles using biological tags and genetic markers. Q5RS-2002-01056 (2003-2006).

WESTHER's overall goal was to describe the population structure of herring stocks distributed from the south-west of Ireland and the Celtic Sea to the northwest of Scotland. To achieve its goal WESTHER had four research objectives: (i) estimation of genetic and phenotypic differentiation between spawning aggregations; (ii) determination of stock origins and life history of juveniles; (iii) determination of composition of feeding aggregations and (iv) improved guidelines for the conservation and management of biodiversity and stock preservation. The Project started officially on January 1 ${ }^{\text {st }}, 2003$ and was extended, in 2005, by six months to finish at the end of June 2006.

A meeting took place in April 2006, of the participants from the different fishery institutes within the project consortium, to discuss the outcomes of each method and their comparisons and to produce a report to fulfil Objective 4: improved guidelines for the conservation and management of biodiversity and stock preservation. At this meeting, four hypotheses were tested and used to inform the deliberations. The null hypothesis was that there is only one herring population to the west of the British Isles, with no detectable differences between any of the geographically and temporally separated spawning components. The following three alternative hypotheses were then tested and discussed. Alternative hypothesis 1: the different spawning aggregations sampled are discrete at spawning time and are, therefore, separate components. Alternative hypothesis 2: there is clear distinction of juveniles sampled on different nursery grounds. Alternative hypothesis 3: fish from each spawning aggregation remain discrete on their feeding grounds. This report was presented to HAWG in 2007 with the recommendations arising from the project's synthesis. The recommendations of WESTHER are given in section 1.3 of this report.

### 1.4.6 The Study Group on Management Strategies [SGMAS]

The Study Group on Management Strategies (SGMAS) met for the third time in January 2007. In previous meetings guidelines have been provided for evaluation of management plans. At this meeting, some plans at various stages of development were revisited, to learn from experience. Furthermore, indicator based management in data poor situations was considered. Finally, the process of developing management strategies, and the role of ICES in such processes was discussed.

The only example stock covered by the HAWG was the Celtic Sea herring. The HCR was a target yield with penalty when $\mathrm{SSB}<\mathrm{Btrig}$, but yield allowed to increase when $\mathrm{SSB}>\mathrm{B}$ trig, both subject to an annual $+-15 \%$ TAC change limit. The experience from that development was that it was not successful. Reasons for that include poor communication between science and stakeholders, problems with the recruitment model (reduced recruitment at the adopted Bpa), and the lack of reliable assessments. Due to the uncertainty in the assessment the approach of using it in the proposed type of HCR will give very conservative yields as the trigger point needs to be well above the point of recruitment impairment.

On indicator based management, the SGMAS considered this to be a promising approach, in particular in data poor situations, but also pointed out that the understanding of the properties and performance of such regimes so far is limited.

The SGMAS emphasized the need for communication and mutual understanding between all interesting parties in the development of management strategies, not the least in the early phase of development. In this phase, the role of science should be to outline opportunities and limitations rather than coming up with specific detailed designs of harvest rules. Later, when evaluating proposed plans, the importance of identifying ambiguities was highlighted, with the recommendation to ask rather than assume.

Several study and working groups have matters relating to the SGMAS work on their agenda. At some stage, there is a need to merge the insight into a unified process for developing and evaluation of management strategies, but so far it is considered more rational to handle different specialized aspects separately. It is clear that the process will have to continue, but at present has not been decided how it will be organized in the future.

### 1.4.7 Workshop on the Integration of Environmental information into fisheries management Strategies WKEFA

Workshop on the Integration of Environmental information into fisheries management Strategies and advice will meet at ICES Copenhagen 18-22 June 2007.

The objective is to identify methodology to operationalize the use of environmental information for the improvement of fisheries management advice. The main thrust of the approach is to take case studies which have consequences for medium term and short term influence in management.

The approach to the workshop has been selected to be compatible with the current annual advice, and the use of management plans as detailed in the report of SGMAS1. The aspects of management advice are considered primarily under single species short term catch options, which follow from medium term management plans based on harvest rates and biomass objectives. Some consideration should also be given to long term implications. This implies evaluation of strategies using criteria of yield, year on year change in yield and the level of risk to the stock under situations of linear or nonlinear environmental change that can influence both the productivity of the stock and the quality of the assessment. For each case study the objective is to identify important life history aspects that change due to environment, including the following

- Recruitment
- Natural mortality
- Growth, Maturation fecundity, including year and cohort effects
- Distribution (habitat and availability)

While it is intended that the main thrust of the meeting will be through the selected case studies, the organisers would welcome detailed proposals for additional case studies, these should be proposed to the organisers as an extended abstract indicating the application of management to be considered and the extent of the effects. The abstract should be submitted no later than 18 May 2007, however, individuals are encouraged to contact the organisers with their intentions as soon as possible,

For each case study the authors need to comment on knowledge and importance, of each of the identified environmental aspects and show how this should influence management and advice.

The Workshop will consider the influence of single or multiple factors on the management of single stocks, where the effects on management are demonstrated, as well as more complex interactions. Authors should select and prioritise the topic areas based on potential influence and available knowledge and should evaluate the impact of change / variability considering;

- Different conditions that influence medium term plans and changes to risk / precautionary limits
- Implications for short term advice and catch options.
- The potential changes in the long term advice and how this might we included in management plans.
- Possibility of extreme events on provision of short term advice.

Where the issues are medium or long term authors should illustrate how this will feed through to both management plans and short-term catch options. Consideration should be given both to management options robust to change as well as reactive management options based on estimation and adaptation.

Specifically the workshop will use the case studies to provide a basis for a synthesis of the needs and roles for management and will provide a report indicating how management advice should be considered, along the lines of environmental influence on:

Short term forecasts
Medium Term management plans
Long term prognosis
The workshop will result in a synthesis report and potentially a paper or collection of papers in a leading journal.

Participants should provide a detailed abstract by 18 May and bring a completed working paper and presentation to the workshop.

NS herring has been selected as a case study we hope to address the

- the influence of spawning stock biomass on recruitment at different environmental conditions and how to determine the contribution?
- the definition of Blim, how to determine it and adjust to different regimes?
- inclusion of recruitment indicators in short-term predictions
- Interannual variability in predation mortality may modify recruiting year classes.
- Appropriate use of year effect and cohort effect growth and maturation
- Inclusion of environmental variables in SRR or adjusting SRR and reference points to productivity regimes,
- Detection and advice during transition phases


### 1.4.8 Workshop on Limit and Target Reference Points [WKREF]

Workshop on Limit and Target Reference Points [WKREF] 29 January to 2 February 2007 in Gdynia, Poland. The TORs were: 1) to review and update the biological basis of limit reference points for fish stocks in the ICES area, taking into account the possible effects of species interactions and regime shifts; 2) to review the scientific and management literature on the implementation of maximum sustainable yield reference points in line with the Johannesburg agreement 2002; and 3) to comment on potential target references points for fish stocks in the ICES area as suggested by SGMAS, taking into account the possible effects of species interactions and regime shifts and the framework on the evaluation of management strategies.

WGREF explored limit reference points for North Sea autumn spawning (NSAS) herring.
In exercises using the segmented regression method (hockey stick) with Norwegian spring spawning herring this approach was found to be quite sensitive to both low S and R values as well as to recruitment values beyond the break point. Due to these theoretical deficiencies alternative methods were explored for NSAS herring.

A simple probabilistic approach to setting $\mathrm{B}_{\mathrm{lim}}$ with the objective to be much less sensitive to recruitment at high biomass than the prevailing SR-methods was presented to the workshop. The method was scrutinised and further developed at the WKREF as a generic approach to stocks that typically have data on low recruitment at low spawning stock size.

The approach focused on low - stock low recruitment where the concept is that below some level of SSB there is an increased probability of a below average recruitment. $\mathrm{P}_{\text {LRi }}$ is defined as the probability of recruitment $\mathrm{R}_{\mathrm{y}}$ (for a number of years y in an ascending sequence of $\mathrm{B}_{\mathrm{y}}$ ) falling below some level $\mathrm{R}_{\text {bar }}$ when spawning biomass $\mathrm{B}_{\mathrm{y}}$ for these years is below some level $\mathrm{B}_{\mathrm{i}}$.
$\mathrm{P}_{\mathrm{LRi}}=\left[\sum \mathrm{y}:\left\{\mathrm{R}_{\mathrm{y}}<\mathrm{R}_{\text {bar }} \wedge \mathrm{B}_{\mathrm{y}}<\mathrm{B}_{\mathrm{i}}\right\}\right] /\left[\sum \mathrm{y}:\left\{\mathrm{B}_{\mathrm{y}}<\mathrm{B}_{\mathrm{i}}\right\}\right]$.
This function is expected to be high at low biomass and be asymptotic to the probability of the level of $R_{\text {bar }}$ for the population. The biomass point $B_{\text {break }}$ at which $P_{\text {LRi }}$ reaches the asymptote is the point where the probability of low recruitment increases.

WKREF considered the probabilistic approach to the entire time series of NSAS herring SRR from $1947-2005$. The breakpoint was evaluated in relation to the probability ( $\mathrm{P}_{\mathrm{LRi}}$ ) of being below the $50 \%$ percentile of recruitment ( $\mathrm{R}_{\text {bar }}$ ). The breakpoint in logistic version was set at $10 \%$ above the estimated asymptotic value. The two model versions gave similar break points $\left(\mathrm{B}_{\text {break }}\right)$ of $0.89 \times 10^{\wedge 6} \mathrm{t}$ and $0.84 \times 10^{\wedge 6} \mathrm{t}$ for the linear and the logistic versions respectively. Model fit to data was slightly higher for the linear version than for the logistic version $\mathrm{R}^{2}=$ 0.982 and $\mathrm{R}^{2}=0.975$ respectively. Residual scatter was approximately normally distributed however some autocorrelation was indicated.

WKREF scrutinised the approach theoretically and concluded that the probability aspect of the method has interesting possibilities because it can specify the probability of obtaining low recruitment. However, the method appears to have some theoretical weaknesses because it does not allow a strict definition of a breakpoint because of the inherent curvature of the probabilistic approach. Further the curvature of probability for low recruitment is dependent on variation in SSR. The method needs further exploration on different types of SRR relationships before it can be applied in an advisory context.

WKREF concluded that there is no basis for changing $\mathrm{B}_{\text {lim }}$ based on current analysis. SGRECVAP results could be basis for revisiting reference point. The distance between a management reference point (trigger or $\mathrm{B}_{\mathrm{pa}}$ ) and $\mathrm{B}_{\mathrm{lim}}$ defines a risk and should be evaluated in the context of harvest control rules in consultation with stakeholders and managers.

## In general WKREF concluded:

that moving to a target F based management would probably remove the importance of $\mathrm{B}_{\mathrm{lim}}$ in a management context.

WKREF has identified three approaches that could be followed in developing long term targets:

1. $\mathrm{EC}\left(\mathrm{B}_{\text {lim }}\right.$ not required; $\mathrm{F}_{0.1}-\mathrm{F}_{\max }$ from yield per recruit analysis as a proxy for $\left.\mathrm{F}_{\mathrm{msy}}\right)$.
2. $\operatorname{ICES}\left(\mathrm{B}_{\text {lim }}\right.$ required, HCR risk analysis: probability of SSB falling below $\mathrm{B}_{\mathrm{lim}}$ )
3. Process/sustainability (NSRAC) (risk analysis also considering socio-economic implications)

For a discussion on advantages and drawbacks of each approach see the WKREF report (ICES CM 2007/ACFM:05).

As indicated above, the question on the role of regime shifts in determining limit reference points was not resolved by WKREF. One approach could be to define different SRR curves for different environmental regimes and to evaluate the breakpoints in these two curves. In general, WKREF recommended looking for biomass limits that would be applicable in both environmental regimes. The distance between $\mathrm{B}_{\mathrm{pa}}$ and $\mathrm{B}_{\mathrm{lim}}$ could take into account the uncertainty due to different regimes.

As HAWG followed this workshop, no extra work was carried out on TOR c other than develop ideas for WKEFA and SGRECVAP.

### 1.4.9 Linking Herring 2008 [ICES/GLOBEC sponsored symposium]

The ICES/Globec sponsored symposium "Herring: Linking biology, ecology and status of populations in the context of changing environments", with the shortened title "Linking Herring" is planned to take place $26-29^{\text {th }}$ August 2008 in Galway, Ireland. The conveners are Maurice Clarke, Mark Dickey-Collas and Aril Slotte. A science organising committee has been set up with Maurice Clarke, Mark Dickey-Collas, Aril Slotte, Emma Hatfield, Doug Hay, Richard Nash, Deirdre Brophy, Øyvind Fiksen as members. The symposium web site is www.linkingherring.com.

The proceedings will be published in the ICES Journal of Marine Science. Niels Daan will act as guest editor on the symposium proceedings.


The Theme Sessions will include:
i) Herring in the middle- the trophic and ecological interactions and impacts of herring
ii Managing Change- management and exploitation of herring in a dynamic environment, within the context of long term change
iii ) Variable Production- particularly the role of reproduction, recruitment and life history strategies.
iv ) Population Integrity- the integrity of stocks and the drivers of migration
v) Counting herring- qualitative and quantitative estimation of herring and its application.

### 1.4.10 Improved advice for the mixed herring stocks in the Skagerrak and Kattegat [EU project IAMHERSKA]

An ecosystem approach to fisheries management should consider conservation of stock and stock sub-component diversity. Spatial variation in composition of stocks or stock subcomponents in areas together with asynchronous population dynamics may lead to overfishing of individual stocks or stock sub-components.

A descriptive analysis of the Danish fleet dynamics during the last decade, in terms of the distribution of herring catches in Division IIIa and Subdivision IVaE, together with an investigation of fleet/metier specific exploitation of the individual stocks in Division IIIa and Subdivision IVaE was performed in the IMHERSKA EU project (Clausen et al., 2006).

Fisheries identified in Ulrich and Andersen (2004) using multivariate analysis of landings profile (target species) and trips descriptors (mesh size, season, and area) were in the IMHERSKA projects modified, to get as much consistency with the previous HAWG work. This resulted in six herring targeting metiers.

The spatial and temporal distribution of the two main stocks (NSAS and WBSS respectively) and the individual life stages (juvenile versus mature) in the Subdivisions IVaE, Division IIIa and Subdivisions 22-24 appear to be following certain patterns in terms of seasonality, which in turn allow spatially and seasonally explicit predictions of the life stage- and stockcomposition in catches. By using the above fleets/metiers and looking at the stock composition in their catches within different areas and seasons, stock selective metiers were identified (a stock selective metier was defined as: a metier with $80 \%$ or more of its landings constituting the same stock).

The ultimate last step of the IMHERSKA project was to bring this data and knowledge together in a metier based projection model, with the potential to predict stock specific Fs depending on how the total catch is distributed between metiers. This projection model is still under development.

### 1.4.11 Study Group on Risk Assessment and Management Advice [SGRAMA]

Whilst the concept of risk is not unfamiliar within ICES and whilst "risk" is commonly understood as the probability of some negative event or harm, most approaches to risk assessment, however, describe risk as consisting of two major components - a probability term and a term that addresses the magnitude of consequence. The need to clarify definitions and terminology as well as to setup a consistent risk assessment and management framework seems obvious. This was the initial intention of ICES to create a new study group in 2006 to deal with this complex topic. Consequently, a new ICES Study Group on Risk Assessment and Management Advice (SGRAMA) met the first time in Copenhagen 18-21 April 2006.

The SGRAMA was created as a first step "in establishing guidelines for production of risk assessments and inclusion of considerations of risk management in the advice. Risk assessment and risk management is considered by ICES as an important field in several branches of science. The SGRAMA aims at drawing on the experience from other branches of science, and to include that experience in the development of risk assessment and risk management in fisheries science." Furthermore, the work of the SGRAMA is considered essential by ICES as "such evaluations are necessary to fulfill the requirements stipulated in the MoUs between ICES and Commissions". Such information is seen to help managers to manage risk in fisheries. As in particular the management component of SGRAMA (the other component is the assessment one) is closely related to the fields of SGMAS, risk management should be considered a part of all management strategies. The assessment part of the SGRAMA should relate to all working groups that are linked to specific stock assessments.

The SGRAMA began its work in 2006 by reviewing different approaches to risk assessment. This limited review discovered a multitude of different use of terminology and definitions. Also because of this, the SGRAMA recommended that the use of the term "risk" should be handled more carefully: "Risk should mean something more than only the probability of some (potentially) harmful event" and "that at least the definition used and the context need to be specified". To tackle this problem, the attempt of SGRAMA was

- to focus on differences in structural approaches and
- to start developing an ICES risk assessment framework
- by setting up definitions
- by concentrating on the clarification of terminology and
- by identifying relevant and important components for it.

This effort is in close compliance with the ICES terms of references a) and b) that are

1. to review and report on available methodologies for risk assessment and frameworks for risk management within and outside the fisheries sector;
2. on the basis of the review, start development of a framework and operational guidelines, for risk assessment and advice which includes considerations on risk management. Risk assessments should inter alia relate to conservation limits and targets for exploitation of fish stocks taking into consideration the ecosystem effects of fisheries and environmental variability and management considerations should relate both to the production of such assessments and institutional aspects of risk management decisions and implementation. The framework should link to the framework for management strategies developed by SGMAS with the scope of ultimately being integrated with these;

The SGRAMA met the second time in Cape Town, South Africa, 5-9 February 2007. This time the aim of the SGRAMA was to consider specific case studies of risk assessment coming from other parts of the world to learn from these. The specific focus this time was on qualitative approaches in risk assessment why the popular "Australian Approach" (Fletcher 2005) was reviewed and discussed. Apart from this, several other working documents were presented, mainly dealing with South African and Namibian experiences in qualitative risk assessment. The only European contribution was a presentation of a quantitative approach regarding risk assessment of North Sea Herring ("Risk assessment of North Sea Herring for stock rebuilding purposes using an optimization algorithm", Gröger 2007). As this approach is closely related to issues of the HAWG it can be considered a first attempt to introduce an integrative approach of risk assessment and an optimization procedure into North Sea herring stock assessments.

### 1.4.12 Workshop on recruitment process of Baltic Sea herring stocks [WKHRPB]

The Workshop on Recruitment Processes of Baltic Sea herring stocks [WKHRPB] was held in Hamburg from 27 February to 2 March 2007 to: conduct a review on recruitment processes of the different Baltic Sea herring stocks; evaluation of the effect of the abiotic and biotic environment of herring recruitment; construction of environmentally-sensitive stockrecruitment relationships; Outline of a scientific project addressing Baltic Sea herring recruitment. The first two tasks were fully addressed while the third was outlined as the possible main TOR in a next year Workshop. Preliminary results of the analysis are presented in section 1.8.

### 1.5 Commercial catch data collation, sampling, and terminology

### 1.5.1 Commercial catch and sampling: data collation and handling Input spreadsheet and initial data processing

Since 1999 (catch data 1998), the working group members have used a spreadsheet to provide all necessary landing and sampling data. The current version used for reporting the 2006 catch data was v1.6.4. All but two nations provided commercial catch data on these spreadsheets, which were then further processed with the SALLOC-application (Patterson et al., 1997). This program gives the needed standard outputs on sampling status and biological parameters. It also clearly documents any decisions made by the species co-ordinators for filling in missing data and raising the catch information of one nation/quarter/area with information from another data set. This allows recalculation of data in the future, or storage and analyses in other tools like InterCatch (see section 1.5.5), choosing the same (subjective) decisions currently made by the WG. Ideally, all data for the various areas should be provided on the standard spreadsheet and processed similarly, resulting in a single output file for all stocks covered by this working group. Two nations failed to deliver their data on time. One of them failed also by the time of the meeting, and still required additional corrections during the meeting, which was rejected.

More information on data handling transparency, data archiving and the current methods for compiling fisheries assessment data are given in the stock annex 2. To facilitate a long-term data storage, the group stores all relevant catch and sampling data in a separate "archive" folder on the ICES network, which is updated annually. This collection is supposed to be kept confidential as it will contain data on misreporting and unallocated catches, and will be available for WG members on request. Table 1.5.1 gives an overview of data available at present, and the source of the data. Members are encouraged to use the latest-version input spreadsheets if the re-entering of catch data is required. Figure 1.5.1 shows the separation of areas applied to data in the archive.

### 1.5.2 Sampling

## Quality of sampling for the whole area.

The level of catch sampling by area in given in the table below for all herring stocks covered by HAWG. The table indicates that the sampling level (in terms of fraction of catch sampled and number of age readings per 1000 t catch) is very different for the various areas. Further details of the sampling quality can be found by stock in the respective sections (Sec. 2.2.4 for North Sea herring, 3.2.6 for Western Baltic Spring Spawners, 4.2.3 for Celtic Sea and VIIj herring, 5.2. for $\mathrm{VIa}(\mathrm{N})$ herring, 6.2 .2 for $\mathrm{VIa}(\mathrm{S})$ and $\mathrm{VIIb}, \mathrm{c}$ herring, 7.2.2 for Irish Sea herring).

|  | Official | Sampled | Aged | Age readings |
| :--- | ---: | ---: | ---: | ---: |
| Area | catch (t) | catch $(\mathrm{t})$ | readings | per 1000 t |
| IVa(E) Summe | 102628 | 89299 | 2784 | 27 |
| IVa(W) Summe | 243561 | 203447 | 15564 | 64 |
| IVb Summe | 92996 | 59479 | 4305 | 46 |
| IVc Summe | 6755 | 2245 | 89 | 13 |
| VIId Summe | 44423 | 35195 | 839 | 19 |
| VIIa(N) | 4402 | 4230 | 1235 | 281 |
| VIa(N) | 27346 | 22135 | 1590 | 58 |
| IIIa | 53172 | 50125 | 12318 | 232 |
| VIIj | 6887 | 6887 | 6501 | 944 |
| VIaS, VIIb,c | 14840 | 14840 | 957 | 64 |

## The EU sampling regime.

HAWG has recommended for years that sampling of commercial catches should be improved for most of the stocks. The EU directive for the collection of fisheries data was implemented in 2002 for all EU member states (Commission Regulation 1639/2001). The provisions in the "data directive" define specific sampling levels. As most of the nations participating in the fisheries on herring assessed here have to obey this data directive, the definitions applicable for herring and the area covered by HAWG are given below:

| Area | SAMPLing level per 1000 t Catch |  |  |
| :--- | :--- | :--- | :--- |
| Baltic area (IIIa (S) and IIIb-c) | 1 sample of which | 100 fish measured and | 50 aged |
| Skagerrak (IIIa (N)) | 1 sample | 100 fish measured | 100 aged |
| North Sea (IV and VIId): | 1 sample | 50 fish measured | 25 aged |
| NE Atlantic and Western Channel ICES sub- | 1 sample | 50 fish measured | 25 aged |
| areas II, V, VI, VII (excluding d) VIII, IX, X, |  |  |  |
| XII, XIV |  |  |  |

There are some exemptions to the above mentioned sampling rules if e.g. landings of a specific EU member states are less than $5 \%$ of the total EU-quota for that particular species.

The process of setting up bilateral agreements for sampling landings into foreign ports has started 2005. However, there is scope for improvement, and more of these agreements have to be negotiated, especially between EU and non-EU countries, to reach a sufficient sampling coverage of these landings.

HAWG reviewed the quality of the overall sampling of herring and sprat for the whole area. There is concern that the present sampling regime may lead to a deterioration of sampling quality, because it does not ensure an appropriate sampling of different metiers (each combination of fleet/nation/area and quarter). Given the diversity of the fleets harvesting most stocks assessed by HAWG, an appropriate spread of sampling effort over the different metiers is more important to the quality of catch at age data than a sufficient overall sampling level. The EU data directive appears to not assure this. The WG therefore recommends that all metiers with substantial catch should be sampled (including by-catches in the industrial fisheries), that catches landed abroad should be sampled, and information on these samples should be made available to the national laboratories.

### 1.5.3 Precision of catch sampling programmes

Port sampling programs aim to provide estimates of the biological composition of the landed catch. Typical characteristics that are collected are length, weight, sex, maturity and age. Of prime interest for stock assessment model input is the catch composition with regard to age. Because age reading is a labour intensive and thus costly method, various sampling schemes have been adopted to make optimal use of age data. Two examples are length-representative sub-sampling and length-stratified sub-sampling. Both methods rely on random selection of individual fish for length determination, but select a non-random subset for age reading. In the length-representative sub-sampling scheme, care is taken to ensure that the randomness at the level of the larger subset carries over to the smaller subset while reducing the probability of introducing bias. The length-stratified sub-sampling scheme makes use of an age-length key, preferably constructed from an independent subset, in order to translate the estimated length distribution into an age distribution. Both of these methods are used within HAWG. Estimates of numbers-at-age in the total landed catch are obtained via multiplication with a raising factor, which can be loosely defined as the inverse of the biomass fraction sampled.

In subsequent paragraphs, several aspects of precision are considered. First, the Swedish port sampling program is described with respect to the variability regarding weights at age. Next, the Dutch port sampling program is described with respect to the precision of numbers-at-age estimates. The Dutch program makes use of a length-representative sub-sampling scheme and
its precision was estimated via an analytical approximation. Finally, precision estimates of the Irish port sampling program are presented. The Irish sampling scheme makes use of agelength keys and estimates were obtained through a bootstrapping procedure.

## Analysis of weight-at-age of Swedish herring IIIa

Sweden has analysed precision levels of calculated CANUM and WECA for several stocks. The results from the analyses are used to establish the sampling plan for 2007.

Here we show the results of the analysis of weight at age (WECA) of Swedish herring IIIa sample in 2006. The replicate in our analysis was the sample vessel ( 50 random individuals). We estimated CV is plotted against the number of fish per age class (Figure 1). Age classes 1 - 4 are included in the graph, which constitute usually more than $90 \%$ of the population in number of individuals. The result shows that CV decrease with increasing number of fish sampled down to a CV of $18-20 \%$. To further decrease the CV, a sampling size much larger than 500 individuals per age class and quarter should be collected. This would increase the sampling costs dramatically (about 4 -fold). Moreover, without a test that is based on such large sample size per age class, it is impossible to predict the number of individuals needed to reach the established CV (established in DCR). For example, based on the relationship in Figure 1, we predicted that with 2006 sample size (around 650 individuals in total per SD and Quarter, more than 150 in the age classes $1-4$ ) we should have reached a CV under $12.5 \%$. However, the observed CV was again around $18-20 \%$. From these results, it was concluded that:

- Any sampling design that is aimed to reach the established CV should be based on experiments with very large and very small sample size to estimate the relationship between N and CV. This would be a very costly procedure but it would elucidate the sample size needed to reach established CV.
- The pattern observed here is likely to be related to the fact that there is an "inbuilt" CV (as it should be expected) in the WECA (as well as in CANUM) and much larger sample size would be necessary to reach the CV aimed by the DCR (i.e. $12.5 \%$ or lower).

It also worth to stress that WECA generates lower CVs than CV calculated for the number at age in catch (CANUM). This would imply that sample size would be even larger for CANUM if established CV should be reached. Thus, in the light of those results, we decided to keep the sampling at planned levels.

The Swedish sampling plan for herring in IIIa are to sample 650 per SD and quarter resulting in about 150 individuals in the age classes 1-4 and a CV around 20\% (see Figure 1.5.3.1). About 50 individuals are sampled randomly from unsorted catches and a total of 10-12 boats are sampled in each quarter and area (Kattegat and Skagerrak), resulting in totally 1300 individuals per quarter in area IIIa.

## Precision of numbers-at-age in the Dutch port sampling program

In the Netherlands, the herring catch is landed in frozen packages of approximately $21-23 \mathrm{~kg}$, consisting of non-sorted fish. According to the Dutch port sampling program, a number of packages of a landing are randomly selected for biological determination. Per package, the number of fish is counted and all are measured for length. A subset of 25 individuals, representative of the length distribution of the package concerned, is selected for assessment of weight, sex, maturity and age. As a result of this length-representative sampling, the 25 individuals are as random a representation of the landing as the selected package is with regard to the length distribution. A schematic outline of the procedure is presented in Figure 1.5.3.2.

Observed numbers are raised to a total per month per area, taking into account the differences in biomass fraction sampled as well as possible spatiotemporal differences in population
composition. Afterwards, the numbers-at-age are summed over areas and over months to arrive at estimates on a quarterly or yearly basis. Alternatively, observed numbers could be raised to quarterly or yearly estimates directly by ignoring spatiotemporal differences on the finer scale. Whether this would result in more precise estimates depends on the data. For the purpose of precision calculations, we have only considered raising procedures on a quarterly basis.

Precision calculations were based on the following basic formula:
[1] $\quad N_{a}=\sum_{s} n_{s} F_{s} P_{a, s}$
Here, $N_{a}$ denotes the total number of fish per age $a, n_{s}$ denotes the total number of fish sampled within a stratum $s, F_{s}$ is a stratum-specific raising factor and $P_{a, s}$ denotes the stratumspecific age probability distribution. Because $F_{s}$ is equal to the aggregate landing weight per stratum $W_{s}$ divided by the product of $n_{s}$ and the average fish weight per stratum $w_{s}$, the above formula can be rewritten as
[2] $\quad N_{a}=\sum_{s} \frac{W_{s} P_{a, s}}{w_{s}}$
This equation illustrates that uncertainty in the total numbers-at-age is governed by uncertainty in the aggregate landing weight and inaccuracy in the estimation of the average fish weight per sampled landing and the sample-specific probability that fish are of a particular age. As the uncertainty in the aggregate landing weight is not due to the sampling program, it will not be considered here.

An exact expression for propagation of estimation errors in $w_{s}$ and $P_{a, s}$ can be obtained if the two stochastic variables can be considered independent. As we did not want to assume independence a priori, a linear approximation was applied to equation [2] by which covariance between $w_{s}$ and $P_{a, s}$ can be taken into account. Estimates of covariance were obtained from variation on the sample level.

Figures 1.5.3.3, 1.5.3.4 and 1.5.3.5 illustrate the method pertaining to raising procedures on a quarterly basis for the years 2004, 2005 and 2006, respectively. Some general characteristics on input data are provided in Table 1.5.3.1. From the figures, it is immediately apparent that standard errors are related to point estimates. However, the relation is not strictly proportional as the relative error is not constant over the age range considered. Relative errors are generally the lowest for numbers-at-age in the third quarter. A striking cohort effect is apparent, in that the lowest relative error is associated with the strong 2000 year-class. Over the age range 2-6 wr, the relative error is generally below $20 \%$. Exceptions are only apparent in the first and final quarter of the year.

## Precision of numbers-at-age in the Irish port sampling program

Irish samples are collected from commercial landings. Length frequency and age data is collected by ICES division by quarter. The length frequency data is added together for each division and quarter and raised to the landings for that area and quarter. The sample weight is divided into the catch weight to get the raising factor. The sum of the length frequencies per quarter is multiplied by the raising factor. An age length key is applied to this data and catch numbers at age calculated.

The precision estimates were worked up using a bootstrap technique. Bootstrapping involves the re-sampling and processing of the source data (measured and aged samples) many times in order to build up a series of results. Precision can then be calculated from the variance observed in the results. For measured data, a sample consisted of a length-frequency distribution. For aged data, a sample consisted of an age and length measurement of a single specimen.

The bootstrap re-sampled with substitution from the collection of measured samples and built up a composite length frequency distribution. For example, if there were five measured samples the algorithm would make five draws from the list with each sample having a $20 \%$ probability of being drawn each time. Re-sampling with substitution from the aged samples in a similar fashion gave an age-length key. Combined with the landings for the stock per quarter the numbers-at-age were derived. After 1000 repetitions, the precision of the numbers-at-age estimate was calculated from the spread of values at each realization. Specifically, the standard deviation of realized estimates was divided by the mean estimate to obtain a relative measure of estimation error.

The results of the method as applied to 2006 data are shown in Table 1.5.3.2 for CS herring and in Table 1.5.3.3 for NW herring. The relative error is below $20 \%$ over the age range 2-6 wr , irrespective of stock. In the third and the fourth quarter, estimates of 1 wr on CS herring were also remarkably precise. At older ages, estimates of NW herring were more precise than estimates of CS herring which is likely due to the higher catch of older fish derived from the NW stock.

### 1.5.4 Terminology

The WG noted that the use of "age", "winter rings" and "rings" still causes confusion outside the group (and sometimes even among WG members). The WG tries to avoid this by consequently using "rings" or "ringers" instead of "age" throughout the report. It should be observed that, for autumn spawning stocks, there is a difference of one year between "age" and "rings". Further elaboration on the rationale behind this can be found in the stock annex 3.

### 1.5.5 Intercatch

"InterCatch is a web-based system for handling fish stock assessment data. National fish stock catches are imported to InterCatch. Stock coordinators then allocate sampled catches to unsampled catches, aggregate to stock level and download the output. The InterCatch stock output can then be used as input for the assessment models." Stock coordinators used InterCatch for the first time at the 2007 Herring Assessment Working Group.

Comparisons between InterCatch and other legacy (previously used) systems were carried out and the maximum discrepancies between the systems are presented in the text table below.

HAWG is the first working group to use this system and some problems were encountered. Ease of use was dependent on the size of the stock and number of allocations required. Allocations can be tedious if dealing with large stocks such as North Sea herring.

The stock coordinators in general found that InterCatch provide a helpful tool at that it has the potential to reduce errors and reduce the work load of the stock coordinators. However several issues should be addressed.

List of suggestions for improvement of InterCatch, some of which are crucial and should be taken very serious:

1) Currently InterCatch does not allow the same full catch information as the Salloc output to be derived. For many stocks in HAWG there is often an issue of misreporting and unallocated catch. Salloc output lists official, working group and unallocated catches separately and, for full transparency, we need InterCatch to do the same.
2 ) The exchange formats used by pelagic working groups currently contain length frequency per quarter, and catch information by statistical rectangle. InterCatch does not include this information. It is recognised by HAWG stock coordinators that this information is not a requirement for assessment input, however, it is very useful in enabling the right allocation decisions to be made and for the
development of length based or spatial models. This information would be a valuable addition to the InterCatch system. It also is a quality control mechanism.
3 ) When uploading and allocating large amounts of information a cross checking procedure would be crucial. A suggestion would be a print version of a list containing all combinations of quarter, area, fleet and country uploaded together with the corresponding CATON. It would then be an easy task for the person responsible for uploading national data into InterCatch to cross-check that all data have been uploaded. Also a cross check list of the allocations made inside InterCatch would be convenient.
$4)$ As it is now InterCatch will not catch the two following types of error: 1) Mismatch between age and CANUM and WECA in the InterCatch input file (e.g. if the data during the copying and pasting from one sheet to another is pasted into a wrong age group simply by mistake). 2) Problems concerning allocation of catches given by rectangle to the wrong areas.
5 ) InterCatch has a security service that prevents data with certain errors to be uploaded and provide easy comprehensible suggestions to where the error is to be found. However, this security system needs further improvement since several type errors in the input files were not discovered by the security system allowing the data to be uploaded but afterwards disappearing. These errors could alternatively be avoided if the check list print version suggested above was available.
6 ) Sprat is caught in vast abundances with numbers often in billions. It is likely this caused the problems encountered during the attempt to upload North Sea sprat data, and this should therefore be investigated.
7 ) Intercatch should be set up to generate some of the standard table (or at least the formatted input for the table) used in the report, otherwise data would in many cases still need to be handled in the ways the respective stock coordinators traditionally have been handling and processing the data.
8 ) As long as the split is not incorporated into InterCatch the stock coordinator of herring in IIIa and SD 22-24 will have to work both with the data in InterCatch and in the traditional way.
9 ) There is currently a lack of authority of stock coordinators to ensure reported data is uploaded to intercatch in the correct format. A formal agreement is needed for institute directors to consent that their staff will do this work.

Maximum discrepancies between InterCatch and other systems:

|  | HER-3A 22 | HER-47 D3 | HER - IRLS | HER - IRLW | HER - NIRS | HER - VIAN |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Caton | $0.01 \%$ | $0.06 \%$ | $0.02 \%$ | $0.01 \%$ | $0.00 \%$ | $0.00 \%$ |
| Canum | $0.01 \%$ | $5.00 \%$ | $0.01 \%$ | $-0.01 \%$ | $0.08 \%$ | $2.17 \%$ |
| Weca | $-7.43 \%$ | $3.30 \%$ | $-0.08 \%$ | $0.13 \%$ | $-0.02 \%$ | $-0.35 \%$ |

### 1.6 Methods Used

### 1.6.1 ICA

"Integrated Catch-at-age Analysis" (ICA: Patterson, 1998; Needle, 2000) combines a statistical separable model of fishing mortality for recent years with a conventional VPA for the more distant past. Population estimates are tuned by CPUE indices from commercial fisheries or research-vessel surveys, which may be age-structured or not as required. This model appears to behave well on the stocks considered by this WG.

The program ICAVIEW4 produces standard plots for the ICA output. However, ICAVIEW does not work on most computers, probably caused by the incompatibility of the program with windows XP and was not used this year. As a result, the standard ICA plots were not presented for all stocks.

### 1.6.2 CSA

"Catch Survey Analysis" (CSA: Mesnil, 2004) is an assessment method that aims to estimate absolute stock abundance, given a time series of catches and of relative abundance indices, typically from research surveys. It does this by filtering measurement error in the latter through a simple two-stage population dynamics model known in the literature as the CollieSissenwine (1983) model. The underlying aim is to reduce the dependence on age-structured data inherent in most VPA-type assessment methods. CSA can be used with only 2 life-history stages (recruits and adults, for example), although simplifying assumptions have to be made. CSA has been used for the exploratory analysis of Celtic Sea herring and North Sea sprat.

### 1.6.3 FLXSA and FLICA [recent developments of XSA and ICA in R]

The FLR (Fisheries Library in R) system (www.flr-project.org) is an attempt to implement a framework for modeling integral fisheries systems including population dynamics, fleet behaviour, stock assessment and management objectives. The stock assessment tools in FLR can also be used on their own in the WG context. The combination of the statistical and graphical tools in R with the stock assessment aids the exploration of input data and results. Currently, an effort is being made to incorporate stock assessment models that are used in some of the ICES working groups. Methods for reading in VPA suite files, for investigating the effect of different model input parameters on the stock estimates, and modeling different aspects of uncertainty are also being developed. Currently the assessment methods "Extended Survivors Analysis" (XSA: Darby \& Flatman, 1994; Shepherd, 1999) and ICA have been incorporated in a package as FLXSA and FLICA, but the development of other stock assessment methods like ADAPT and SURBA is on-going.

During this year's assessment, the FLICA package was adjusted to provide raw parameter estimates together with the variance-covariance matrix as standard output from ICA. With this information, the standard diagnostics of ICA were replaced with diagnostics generated within FLR. The WG decided to show results of catchability models and regression residuals as they are actually fitted. Thus, observed indices are treated as dependent variables and VPA estimates of SSB or numbers at age are considered predictor variables. This enhances the visual judgment of the quality of model fit, even though the nature of the data would suggest a reversal of predictor and dependent variables. It may be sensible to take this into account in the way the catchability models are fitted, but this would require changes in the ICA code itself. In addition, two plots were added to the diagnostic output: a Q-Q plot to show the distribution of the log residuals as compared to a normal distribution; and an autocorrelogram to show the autocorrelation function of the log- residuals. These two plots are shown because the catchability models fitted assume a normal distribution of the $\log$ residuals and no autocorrelation therein.

In this working group, FLR has been used for exploratory analyses of North Sea herring (FLICA for deterministic and retrospective analyses), herring in IIIa, Celtic Sea (exploratory analysis) and Via South.

### 1.6.4 SURBA

"Survey Based Assessment" (SURBA: Beare, 2005; Needle, 2003, 2004) is based on a simple survey-based separable model of mortality. At the moment SURBA is not yet available in FLR, but development towards this is ongoing. While SURBA was used in the past in this WG it was not implemented this year.

### 1.6.5 MFSP, MSYPR and MFDP

Short-term predictions for the North Sea used MFSP / MSYPR that was developed three years ago in the HAWG (Skagen; WD to HAWG 2003). Other short-term predictions were carried out using the MFDP v.1a software.

### 1.6.6 STPR used for medium term projections NS herring

Medium term projections were performed with the STPR3 software, supplemented with a version (S3S) made to ease screening over ranges of model parameter choices. The software documentation is available from ICES or as a report (Skagen, 2003). The simulation framework covers alternative scenarios for future recruitment, weight and maturity at age, assessment error, discarding and other unaccounted mortality. The harvest rules can be examined with respect to error in future assessments by assuming that the stock numbers at age, and hence the SSB on which managers make their decisions, deviates from the real state of the stock. STPR3 does this by a simple stochastic multiplier on the stock numbers as seen by decision makers. Likewise, discrepancy between the decided TAC and the catch actually taken is simulated by a common implementation multiplier. This may account for bias due to misreporting etc. Uncertainty due to measurement (i.e. sampling of the catch derivation of CPUE) estimation within the assessment process, model mis-specification and implementation error were not explicitly modelled but assigned a combined assessment error. However, varying feedback between the assessment process and the management decision making process was not included. Feedback can cause bias in the assessment to affect the management and thus the stock which in turn affects bias in the assessment.

The simple approach in STPR allows for some evaluation of the robustness of a harvest rule to such errors, but does not pretend to foresee how these errors will appear in the future. However, to be feasible, one would assume that the harvest rule still should lead to a precautionary management if these errors have an order of magnitude that has been experienced in the past. It may be noted that previous implementation error that has not been accounted for, although it will have influenced the perception of the stock in the past. Hence, implementation error should only cover cases where it may be different from what it was in the past or already documented and explicitly included in past data.

### 1.6.7 Management simulations

In order to evaluate the impact of alternative scenarios of stock and recruitment for the North Sea herring stock population dynamics, an evaluation platform has been implemented, including four model components. In order to develop the platform, an age-structured population dynamics model has been developed (McAllister, Pikitch et al. 1994; Punt, Smith et al. 2002). The model allows a realistic representation of the population dynamics taking into account potential bias in observations through the observation error model (including all surveys available). The model also allows to evaluate the state of the stock using ICA as the assessment method and to utilise the actual management procedure to provide management advice using the harvest control rules model. The simulation-testing framework has been developed and implemented under FLR using several packages such as FLICA, already used last year by the WG. This platform was not used to evaluate the existing HCR yet, but the approach is being developed in the EU project FISBOAT (http://www.ifremer.fr/drvecohal/fisboat). During this year's meeting, substantial input was provided to the developer by HAWG.

### 1.6.8 Bayesian Statistical Catch-at-age

An exploratory analysis of the Celtic Sea herring data was performed by means of a statistical catch-at-age model. The model uses Bayesian estimation and was implemented in WINBUGS
(Spiegelhalter 2003). The statistical catch-at-age model was used for the period where survey data is available (1995 to 2006). The early part of the series is derived from a VPA with starting numbers from 1997. The fishery was fitted with a logistic selection function that can change from year to year. The example shown to the WG allowed only a slow change in selection.

The model exploration was considered preliminary, there was no evaluation of the influence of priors, though they were thought to be uninformative, and only a limited range of flexible selections were tested.

### 1.7 Discarding and unaccounted mortality by Pelagic fishing Vessels

In many fisheries, fish, invertebrates and other animals are caught as by-catch and returned to the sea, a practice known as discarding. Most animals do not survive this procedure. Reasons for discarding are various and usually have economic drivers :

- Fish smaller than the minimum landing size
- Quota for this specific species has already been taken
- Fish of undesired quality (high-grading) or low market value
- By-caught species of no commercial value.

Theoretically, the use of modern fish finding technology used to find schools of fish should result in low by-catch. However, if species mixing occurs in pelagic schools (most notable of herring and mackerel), non-target species might be discarded. Releasing unwanted catch from the net (slipping) or pumping unsorted catch overboard also results in discarding.

In the area considered by HAWG, only 3 nations reported discards from their fleets in 2006. From those, only two incorporated discards in the assessment data.. The discard figures were raised to national landings (based on the spatial and temporal distribution of the fleet), and used in the assessment of North Sea autumn spawning herring (UK/Scotland and Germany, see Section 2.3) and VIaN (UK/Scotland, see Section 5.1.3). For the Netherlands, the estimates of discards of approximately 4 thousands tonnes per year were not sampled at a high enough resolution to allocate the catch in individual stocks.

All other nations did not report notable amounts of discards of herring in the pelagic fisheries, either because they did not occur, catches were not sampled for discards or there were difficulties with raising procedures (ICES, 2007/ACFM:06). No discard estimates for the total international catch were calculated, on a basis that some of coverage is still not high enough.

Very few estimates of discarding of pelagic species from pelagic and demersal fisheries have been published. Discard percentages of pelagic species from demersal fisheries were estimated between $3 \%$ to $7 \%$ (Borges et al., 2005) of the total catch in weight, while from pelagic fisheries were estimated between $4 \%$ to $11 \%$ (Pierce et al., 2002; Hofstede and Dickey-Collas, 2006). Even less information exists on the discarding of non-commercial fish.

For the Dutch pelagic freezer-trawler fleet, a first ever estimation of discarding was carried out based on observers on-board vessels (Borges et al, working document). A total of 27 trips and 904 hauls were sampled between 2002 and 2005, covering the North Sea and western waters of the British Isles. This study intended to investigate suitable methods for estimating discarded fish by the Dutch pelagic freezer-trawler fleet. This fleet is viewed as fairly typical of similar vessels that operate under German, UK and French flags, which are mostly Dutch owned. Different methods to estimate the total discards were compared and raising observations by trip or by landing did not affect greatly the annual estimates of total discarding, except for 2003.

The results show that for this fleet which has high volumes of catches of a few species (mainly herring, mackerel, horse mackerel and blue whiting), approximately $10 \%$ of the total catch discarded is discarded, showing that it is a selective fishery for its targeted species. However the majority of discards are from not targeted. The percentage of catch discarded per species shows mackerel highly discarded ( $35 \%$ of the total catch) and this can be caused by either quota limitation or landings misreporting. For the horse mackerel and the herring, discards are respectively $7 \%$ and $5 \%$ of the total catch. Estimate discards par year for herring of approximately 5500 tonnes (Figure 1.7.1) are low when compared to total landings of the Dutch pelagic freezer-trawler fleet (Figure 1.7.2). These estimates include slippages.

The inclusion of discarded catch is considered to reduce bias of the assessment and thus give more realistic values of fishing mortality and biomass. However, they might also increase the noise in the assessment because the sampling level for discards is usually lower than that for landings. This low sampling rate is caused by the large number of different metiers in the pelagic fishery and the difficulty of predicting behaviour of the fisheries (in terms of target species and spatial and temporal distribution). Raising discard estimates to the national landings might result in a higher bias than an area based estimate of discards from the total international fleet, if sampling is insufficient. HAWG therefore recommends that the development of methods for estimating discards be based on a fleet based method, rather than on a national basis. Table 1.7.1. and 1.7.2 show the number of samples done in 2006 for the pelagic fleets by country.

### 1.8 Ecosystem considerations, sprat and herring- response to WGRED, SGRECVAP and SGRESP.

HAWG acknowledges the significance of the variability in the ecosystem as an important driver of the herring and sprat stock dynamic. This must be considered when giving advice. Despite the increasing pressure on working groups to consider their allocated stocks within the context of the ecosystem, the potential added value of having targeted ecosystem groups (such as NORSEPP, WGRED and REGSNS) is still minimal due to the lack of an interaction between these groups and the assessment groups. The provision of the data by the ecosystem groups and the summaries they provide are still largely unsuitable for consideration and adoption by assessment working groups. This is partly due to their acting in isolation. Although assessment working groups are generally populated by scientists with a "stock assessment" slant, HAWG has a history of using and investigating environmental drivers and changes in productivity, and such work has fed into and been used by groups such as SGPRISM, SGRESP, SPACC and other GLOBEC groups. Summaries of physical and environmental times series that reflect the dynamics of the NE Atlantic and environs of the North Sea are required by HAWG. These summaries must be cumulative and not "stand alone" quarterly reports, and they should document variability and fluctuations of inflow, transport, primary and secondary production, water column stability, turbulence, salinity and temperature.

Examples of the use and interest of HAWG in the dynamics of the ecosystem and its impact of the fish stocks include:

- the use of shifts in recruitment productivity in North Sea herring in short and medium term projections
- the accounting for productivity changes in the development of management scenarios for west of Scotland herring
- the analysis of surplus and net productivity in herring stocks in relation to fishing mortality
- by incorporating, whenever possible, empirically derived annually variable weights and maturity ogives in stock assessments
- by accounting for cohort specific dramatic changes in weight and maturity in short term projections
- investigations of the dynamics of the timing of spawning and the temporal origin of fish in both the catch and surveys
- the investigations of the between year larval mortality in North Sea herring
- investigations of fecundity in herring
- the search for more robust indices of recruitment in all stocks
- initiating work on the interactions of multispecies catches of the fleets that target small pelagics

HAWG welcomes that North Sea herring will be a case study in the work shop on Workshop on the Integration of Environmental Information into Fisheries Management Strategies and Advice (WKEFA).

### 1.8.1 North Sea

Salinity and temperature are known to have a large impact shaping the ecosystem structure in the North Sea and generally their variability reflects the influence of the North Atlantic Oscillation (NAO) on the movement of Atlantic water into the North Sea. The long-term temperature and salinity anomalies in the Atlantic waters flowing into the North Sea with the Faroe Isle current provide a broadly similar cyclical behaviour up to the late 1990s (ICES 2006/LRC:03). However, in more recent years the two signals appear to diverge, with relatively high temperatures persisting during years showing a marked decline in salinity (Hughes and Lavin 2005).

In 2005, SST (Sea Surface Temperature) was close to the long term mean for the first eight months of the year, but showed strong positive anomalies in September to December (source http://www.bsh.de/en/index.jsp). The last quarter of 2006 was characterised by much warmer condition ( $1-3 \mathrm{C}^{\circ}$ in the eastern part and about $1 \mathrm{C}^{\circ}$ in northern part of the North Sea) when compared to the average of the last 50 years (NORSEPP 2006).

A series of studies on the plankton ecosystem at the herring spawning grounds in the North Sea have shown a strong linkage between frontal hydrography and the prey availability, growth and drift of herring larvae which hatch in these areas (e.g. Richardson and Heath 1986, Kiørboe et al 1988) and other studies propose a strong connection between frontal hydrography and herring recruitment (Iles and Sinclair 1982, Sinclair 1988). Preliminary explorations of the hydrographic variability at the spawning grounds during the period 19752005 indicate that two periods of poor herring recruitment (1987-90 and 2001-05) coincide with periods of anomalous low water density in nearshore areas. This observed decline in water density is connected to both salinity and temperature fluctuations, and has most likely changed the performance of nearshore fronts. Hence, the preliminary comparisons indicate that the herring recruitment could be affected by oceanographic fluctuations, leading to changes in frontal hydrography at the herring spawning areas, and further investigations of this relationship are recommended.

In concomitance with an increased SST and decreased salinity, the plankton community in the North Sea has shifted to a dominance of more "southerly" species, as shown by CPR data (Reid et al., 2003) after the decline in the abundance of the copepod Calanus finmarchicus. Both Calanus species and juvenile sand eels are common prey of herring and recent evidence from the Baltic has shown that herring positively select Pseudocalanus and Temora and select less Acartia (Casini et al., 2004). Acartia is associated with summer blooms and warmer temperatures as shown by Gowen et al. (1998). These trends in zooplankton species abundance and species compositions appear to be continuing and those might have causal effect with herring growth and migration patterns (ICES 2006/ACE:03). The CPR data also show a reduction in euphausid availability. Although no changes have been recorded in the
total zooplankton biomass and in total copepod abundance (e.g. northern North Sea; SAHFOS 2004, Heath 2005), the overall picture is one of a changing zooplankton community structure.

The production of herring has increased (ICES 2005/ACFM:18) since the collapse caused by overfishing in the 1970s (for methods details see Dutil and Brander, 2003). Surplus production has been of the order of 700 k tonnes for the last 25 years and the recent positive net production has lead to an increase in available herring biomass in the system. Also, the biomass of sprat is considered high and fairly stable compared to the last decade (ICES 2005/ACFM:18) (see also section 1.8.3).

In terms of the impact of a high biomass of herring and sprat on the North Sea ecosystem, some studies are ongoing, but more resources are required to obtain new estimates of stomach contents, prey selectivity, stomach evacuation rates and behavioural interactions by herring and sprat. With low sandeel and Calanus abundances, the herring may well be having a stronger impact on the ecosystem than in the previous last 2 decades. However, a high biomass of herring may also provide an alternative prey source to piscivores such as horse mackerel and Minke whales (Olsen and Holst, 2001) reducing the pressure on sandeel. Also, the impact of herring as predator of fish eggs varies with the prey spectra faced by the species (Segers et al., 2006). These last sentences are very speculative and if the quantitative trophiccomplexities of the system are to be considered as a priority by ICES, more resources need to be spent on understanding the trophic interactions in the North Sea and developing spatial and temporal trophic dynamics models of the system.

Recent investigations of the decline in larval herring at age (empirical data from the ICES coordinated larval herring surveys, Dickey-Collas in prep.), which used a temperature dependent growth model to estimate larval age, suggest that the daily mortality rate of herring in the North Sea has recently increased to the highest in the time series (ICES 2006/LRC:03). There was evidence that high mortality of herring larvae can co-occur with high larval production (CM 2006/LRC:03). There was a strong negative trend in the residuals from stockrecruits relationship in the latest decade suggesting that the poor recruitment is not just related to high spawning stock biomass level (ICES 2006/LRC:03) but likely caused by an high mortality of herring larvae. The mechanisms for this were most likely poor larval feeding, predation, poor hatching condition and probably a combination of those with possible links to variable hydrographic conditions.

SGRECVAP (ICES 2006/LRC:03), using dynamic factor analysis, highlighted a positive correlation between the time series of SST and herring recruitment anomalies in the North Sea. In addition, a recent analysis (see Cardinale and Hjelm, 2006 for details on methods used) on the effect of spawning stock biomass (SSB) and sea surface temperature (SST) on clupeid recruitment in the North East Atlantic showed significantly more stocks with an SSB effect on recruitment compared to an SST effect on recruitment (Cardinale et al., 2006), although there was not significant difference on the strength of the SSB compared to the SST effect. Variability of recruitment anomalies of clupeid stocks (using all stocks assessed by ICES in the North East Atlantic) was positively strongly correlated with anomalies of SST in the area. A strong positive relationship was found between the first principal component, which explained around $29 \%$ of the recruitment anomalies variation, and average temperature deviations in the area. A similar relationship was found also with NAO, but its strength was lower than for SST. Interestingly, 70s and 80s are clearly separated from 90s and onwards, plausibly mimicking the different climate regimes (i.e cold against warm period) (Cardinale et al., 2006). This again highlights the link of temperature to recruitment strength of clupeids in the area but it does not provide any clear underlying mechanisms.

Recent analysis of the occurrence of sardine and anchovy as recorded in IBTS survey has increased rapidly in the last $10-15$ years in the North Sea (Figure 1.8.1.1). Those are species adapted to warmer oceanic conditions and we could speculate that their increase is linked with
warming of the North Sea. However, this investigation is preliminary and a spatial analysis of the occurrence of those species in the North Sea would add insight on the observed phenomenon.

The Kattegat and the Skagerrak is also considered an important area for herring by HAWG, it supports both local spawning populations and is the major nursery ground for North Sea herring. The impact of the higher saline inflows through this area into the Baltic Sea in recent years on the resident herring populations is at present unknown. Studies presented to HAWG in 2005 about the HERGEN (Bekkevold et al., 2005) project suggest that salinity may play a role in the genetic integrity of local spawning components. A preliminary analysis made at the WKHRPB (ICES 2007) was unable to find any climate signal on the recruitment of Western Baltic herring although the time series is rather short. For all other pelagic fish stocks in the Baltic area, a positive effect of a warm regime was evident although spawning stock biomass seems to play a more critical role for regulating recruitment in most of the stocks with the exception for stocks distributed in the gulfs (i.e. Gulf of Riga herring and Bothnian Bay herring) (ICES 2007).

In the neighbouring Baltic Sea, the interactions between herring and sprat have been shown to be very dynamic (Mollmann and Koster, 2002). A close association in food items predated upon by those species has been recently shown, together with a clear density-dependent (i.e. food limited) growth for both herring and sprat (Casini et al., 2006). Clupeid condition covaried with the changes in the weight of zooplankton in the stomachs, which further suggest food competition being the main mechanism behind the changes in clupeid condition in the Baltic Sea during the last two decades. This is the first evidence of food resource mediated density-dependent fish growth in a large marine ecosystem (Casini et al., 2006). The individual fish from the strong 2000 year class of herring in the North Sea have been smaller in size and are less mature at age. This suggests that either slower-growing fish have survived in that year class or that the ecosystem has failed to provide enough food to allow the full potential growth for that cohort i.e. that food has been limiting for that cohort. This cohort grew well up to 1 winter ring of age. However, the less abundant 2001-2004 year classes show again average growth, tending to corroborate food limitation as the likely explaining factor for growth rates variability also in the North Sea herring (ICES 2006/LRC:03). With the decline in sandeel and other planktivorous fish, HAWG would support further studies into the feeding interaction and spatial and temporal associations of herring, sprat, anchovy and pilchard (sardine), especially in the light of the increase of the abundance of the latter southern species in the area during the latest decade (ICES 2006/ACE:03).

Most herring fisheries deploy gear that is deployed clear of the seabed. The impact of gravel extraction on the conservation and productivity of herring is still unclear, and there are virtually no studies to provide evidence at present (ICES 2005/ACFM:18). The limited evidence available at present records no incidences of cetacean mortality due to pelagic trawling ( 0 catches observed out of 218 pelagic hauls by commercial trawlers from 19992004). There are also very few other by-catches of fish, beyond the targeted fisheries of herring, mackerel, horse mackerel and blue whiting.

No specific environmental signals were identified specifically by WGRED (CM 2006/LRC:03) to be considered in assessment or management of herring and sprat in this area in 2006.

A possible link between ecosystem changes and the dynamic of North Sea herring might be the yearly variation of age specific natural mortality. HAWG (2007) has evaluated the effect of variable $M$ (derived from MSVPA) on the historical dynamic of North Sea herring. Estimates of SSB made using variable $M$ were similar to those using fixed $M$ (used in the current assessment). However, and as expected, the use of variable M affected the perception of recruitment and F of the juveniles (age 0-1). Recruitment was smaller except for 1990 to

1995 year classes, while F (0-1) was slightly larger than with fixed M. The use of variable M also changed the relationship between stock and recruitment, making it more similar to a Beverton and Holt curve than previously believed. However, the largest changes appear for the segmented regression where the SSB Break point is largely reduced using fixed M, although the use of the segmented regression for estimating Blim has been recently questioned (WKREF 2007). (Figure 1.8.1.2).

### 1.8.2 Celtic Seas

The western herring stocks assessed by HAWG are found in the Celtic Seas (Celtic Sea, Irish Sea, Malin/Hebridean Shelf). There is less information on the hydrographic variability and ecosystem dynamics in the Celtic Seas. WGRED appeared to concentrate on the Celtic Sea.

## Celtic Sea

In the Celtic Sea, in terms of hydrographic variability, the Irish Shelf Front, that occurs to the south and west of Ireland (at about $11^{\circ} \mathrm{W}$ ) around the 150 m isobath, and exists year-round, is an important feature for the structure of the marine ecosystem in the area. The turbulence caused by the front may bring nutrients from deeper water to the surface where it promotes the growth of phytoplankton, especially diatoms in spring, but also dinoflagellates where there is increased stratification. These may in-turn be fed on by swarms of zooplankton and associated with these, aggregations of fish, like herring and sprat (Reid et al. 2003).

The WGRED report (ICES 2006/ACE:03) suggests that are indications of steady warming in the area over recent years. Similar trends appear for salinity (ICES 2006/ACE:03). Considering that Celtic Sea herring is the second most southerly population of herring exploited in Europe, and this is an area of warming sea surface water, sea warming could affect the recruitment of this pelagic species.

Variation of zooplankton abundance and species composition might affect feeding conditions and mortality of juveniles and adults of both herring and sprat. Zooplankton monitoring data are available from one station in waters about 50 m deep in the English Channel. These data exhibited a decreasing trend from 1988 to 1995 but a recovery thereafter. This recovery was mainly due to two autumn developing small species of copepod, Euterpina sp. and Oncaea sp. In 1999 there was a decline in the zooplankton population, with the top ten species all below their typical average values (apart from Temora and Corycaeus, which exhibited very little variation) (ICES 2006/ACE:03). In 2000, 2001 and 2002 zooplankton population abundance experienced a recovery reaching values comparable to those after 1995 (reported in ICES Zooplankton Monitoring Status Summary 2001/2002). Data for 2004, 2005 and 2006 were not yet available.

WGRED considered that in the Celtic Sea a key pelagic species here is herring as well as sardine, in the southern area, and sprat, in the Celtic Sea proper. The area also accommodates considerable stocks of argentines (two species) and large numbers of small mesopelagic myctophids along the shelf break (ICES 2006/ACE:03).

Southward et al (1988) demonstrated that the abundance of herring Clupea harengus and pilchard Sardina pilchardus occurring off the south-west of England closely corresponded with fluctuations in water temperature. Sardine was generally more abundant and extended further to the east when climate was warmer whilst herring were generally more abundant in cooler times. This pattern has apparently been occurring for at least 400 years, and major changes were noted in the late 1960s as waters cooled and spawning of sardine was inhibited. In recent years herring populations have declined throughout the Celtic Seas ecoregion but are unclear whether sardine have increased in abundance.

Despite recent evidence from WESTHER and HERGEN that there is little genetic differentiation between herring stocks, their phenotypic characteristics and population dynamics are different. The Celtic Sea shows a very different pattern compared to both the west of Scotland and the Irish Sea stock (ICES 2006/ACE:03) (see section 1.8.3).

No obvious environmental signals were identified by WGRED that should be considered in assessment or management of herring and sprat in those areas. However, the major trends in the ecosystem noted above (i.e. the steady warming of the area and the reduction of copepod abundance) could play a major role to shape the dynamic of herring and sprat stocks in the near future (ICES 2006/ACE:03).

### 1.8.3 Investigating Productivity

The North Sea herring is a long and well documented species in terms of its exploitation and related collapses and recoveries (Cushing and Bridger 1966, Burd 1985, Nichols 2001, Simmonds 2005). The main impact on its productivity was generally expected to be fishing, although the environment may have a major impact as well. Each habitat or ecosystem is assumed to have a carrying capacity which varies in time (Jennings et al. 2003). To account for the influence of the ecosystem on the productivity of five different herring stocks (Tab. 1.8.3.1) two different methods were applied (Nash and Dickey-Collas 2005).

First the recruit per spawner ratio was calculated. High ratios were assumed to represent a high production and low ratios a low production. These calculations formed the basis for the detection of periods of high and low production of the stock.

The next step was to calculate the net and surplus production of the whole stock, including the recruits and the growth of all non-recruits, the natural and the fishing mortality. To subtract the influence of the spawning stock biomass a hockey stick and a Ricker stock recruitment relationship were fitted to the data to obtain the residuals of the recruits of a given year. The residuals were used to remove the year effect from the estimation of the stock size and to gain the net production and the surplus production respectively without the effect of the SSB on the number of recruits. The parameters used to fit the data from the different stocks are given in Table 1.8.3.2.

The data used in this analysis was derived from the assessment outputs from the HAWG in 2006 (Table 1.8.3.1). All stocks the HAWG dealt with are used, except the Western Baltic spring spawning herring (IIIa herring). The time series of the IIIa herring was assumed to be too short to meet the requirements of the analysis used.

Calculation of the surplus production
$\mathrm{Ps}=\mathrm{Br}+\mathrm{Bg}-\mathrm{M}$
where Br is the biomass of the recruits, Bg the gain of biomass due to growth of all fish excluding the recruits and M the natural mortality. The net production equals the surplus production minus the fishing mortality $(\mathrm{F})$.

The impact of a varying F was tested using the North Sea herring time series as an example with both stock recruitment relationships adapted to the dataset.

All stocks showed highly variable production over time (Figures 1.8.3.1 and 1.8.3.2) that can be seen both in the recruit per spawner as well as in the net and surplus production estimates derived from the calculations that take the year effect into account. Except the Celtic Sea herring, all stocks showed markedly changes in the average productivity between different periods. However, these periods are not synchronized between the stocks. In the North Sea the productivity increased markedly after the collapse in the late 70 's, supporting the recovery of the stock. In the middle of the 80 's the productivity fell to the level before the collapse. In the
last three years the productivity fell again (Figure 1.8.3.1). It is assumed, that without decreasing the fishing impact on this stock a collapse will happen again. If the recovery will be again supported by a high production due to good environmental conditions is doubtful.

The North Sea stock was also chosen to show the influence of different $F$ on the outcome of the two fitted stock recruitment relationships. The overall trend is a decrease both in net and in surplus production with increasing F. Nevertheless the general pattern is still conserved. Therefore, for all other stocks only the figures derived with an F of 0.25 are chosen.

The Irish Sea herring stock showed a marked decline in productivity during the late 70's and remained at a low level since then. This feature is represented in the recruit per spawner ratio as well as in the net and surplus production (Figure 1.8.3.2a). The Celtic Sea herring stock had a low productivity throughout the whole time series. However, the net and surplus production is very noisy displaying neither clear trend nor a general low productivity (Figure 1.8.3.2b). The $\mathrm{VIa}(\mathrm{N})$ herring stock showed a variable recruit to stock ratio without marked periods. However, the net and surplus production seemed to present a period of high productivity in average from the 60 's to the 90 's and a lower average in the recent decade (Figure 1.8.3.2c). $\mathrm{The} \mathrm{VIa}(\mathrm{S})$ herring stock time series is shorter than the others. Therefore, general trends were not apparent. Nevertheless a high production in the 70 's and 80 's was followed by a sharp decline and a slow increase in the 90 's (Figure 1.8.3.2d).

### 1.9 Pelagic Regional Advisory Council [Pelagic RAC]

Members of HAWG have attended meetings of the pelagic RAC since its inauguration in 2005 and throughout 2006 and into 2007. HAWG considers the views of the Pelagic RAC as important, and welcomes the formation of the forum to give stakeholders a role in the advisory process. HAWG notes that the Pelagic RAC also has special members from outside the EU, notably from Norway.

Most relevant documents from the Pelagic RAC to ICES and the European Commission about herring assessment and management were available to HAWG through the meeting.

### 1.10 Stock overview

Analytical assessment could be carried out for three of these eleven stocks. Results of the assessments are presented in the subsequent sections of the report and are summarized below and in Figures 1.10.1-1.10.3.

North Sea autumn spawning herring is the largest stock assessed by this WG. It has experienced very low spawning stock biomass levels in the late 1970s when the fishery was closed for a number of years. This stock began to recover until the mid-1990s, when it appeared to decrease again rapidly. A management scheme was adopted to halt this decline. Following a period of good recruitment co-occurring with the new management measures, SSB and the proportion of older fish in the stock increased. This gave the opportunity to increase TACs and catch. The recent trends in SSB show that after a peak of 1.8 million tonnes in 2004, the SSB in 2006 was 1.2 million tonnes. The current fishing mortality $\left(\mathrm{F}_{2-6}\right)$ is 0.35 and is well above the target F prescribed the management agreement. It is likely that the stock will decline further close to $\mathrm{B}_{\text {lim }}$ by 2008. The decline in SSB is due to serial poor recruitment since 2001 and a failure to fish at target F for the adults in the last few years. The estimate for the most recent recruiting year class is the lowest since 1979 and the low recruitment is caused during the larvae phase of the North Sea herring.

Western Baltic Spring Spawners (WBSS) is the only spring spawning stock assessed within this WG. It is distributed in the eastern part of the North Sea, the Skagerrak, the Kattegat and the Sub-Divisions 22, 23 and 24. Within the northern area, the stock mixes with North Sea autumn spawners. An analytical assessment demonstrates that SSB has been slightly
increasing or stable over a number of years. When compared to possible MSY target fishing mortalities, it is likely that the current fishing mortality is too high. There is an indication of a declining recruitment in recent years in the WBSS herring stock.

Celtic Sea herring: The herring fisheries to the south of Ireland in the Celtic Sea and in Division VIIj have been considered to exploit the same stock. For the purpose of stock assessment and management, these areas have been combined since 1982. The fishery in the eastern part of the Celtic Sea was closed in the early eighties due to poor recruitment. In 2007, HAWG carried out a benchmark assessment on the Celtic Sea herring. The exploration showed that there is uncertainty in $\mathrm{SSB}, \mathrm{F}$ and recruitment for last 3 years. However, information from the catch shows an increasing trend in the mortality of the fish and a contraction in age structure of the stock. Exploration with simpler models showed a decline in biomass over the whole last 10 years. A Bayesian analysis suggests that the selection of the fishery has changed over the last 10 years, and supported the perception that the current status of SSB is uncertain, but probably at a low level. Analysis of recruitment patterns suggested that no major regime shift has taken place in Celtic Sea herring productivity in the last 40 years.

West of Scotland herring was recently regarded as lightly exploited, but in 2006, the stock was more heavily exploited than it has been since 1999 . Earlier data indicate the possibility of larger stock in the 1960s when the productivity of the stock was different from now. The stock experienced a heavy fishery in the mid-70s following closure of the North Sea fishery. The fishery was closed before the stock collapsed. It was opened again along with the North Sea. In the mid 1990s there was substantial area misreporting of catch into this area and sampling of catch deteriorated. Area misreporting was reduced to a very low level and information on catch has improved, but in 2004 and 2005 misreporting increased again. In 2006, however, there was no misreporting from IVa into VIa (N). In 2006 the dominant year classes were 1999 and 2000. It appears that the 2001 year class is not strong as was originally supposed, but relatively weak. Recruitment seems to be low since 2001, but the level of recruitment at 1wr in 2006 and 2007 is uncertain.

Herring in VIa south and VIIbc are considered to consist of a mixture of autumn- and winter/spring-spawning fish. The winter/spring-spawning component is distributed in the northern part of the area. The main decline in the overall stock since 1998 appears to have taken place on the autumn-spawning component, and this is particularly evident on the traditional spawning grounds in VIIb. The current levels of SSB and F are not precisely known, as there is no tuned assessment available for this stock. There are no sign of stock recovery in VIaS herring.

Irish Sea autumn spawning herring as comprises of two spawning groups (Manx and Mourne). This stock complex experienced a very low biomass level in the late 1970s with an increase in the mid-1980s after the introduction of quotas. The stock then declined from the late 1980s to its present level. During this time period the contribution of the Mourne spawning component has declined. In the past decade there have been problems in assessing the stock. It seems likely that the stock has been relatively stable for the last 10 years, and that the fishing mortality does not appear to be increasing above the recent average. The catches have been low in recent years and the fishing activity has not varied considerably. There is evidence of a contraction in the age structure of this stock. Recruitment is approximately average for the period since the 1980s.

North Sea Sprat is the only sprat stock on which an assessment is carried out within this WG. Sprat in the North Sea is a short-lived species. The recruits account for a large proportion of the stock, and the fishery in a given year is very dependent on that year's incoming year class. The size of the stock has been variable with a large biomass in the early 90 's followed by a
sharp decline. It is likely that the abundance of North Sea sprat is now less than in the last two previous years.

### 1.11 Structure of the report

The report below, further details in each chapter the available information on the catch, fisheries and biology of the stocks and then the stock assessments, the projections, the quality of the assessments and management considerations for each stock. This information and analysis are given in chapters for each of the seven major stocks considered by HAWG. Despite this structure, it is important to realise that there are many links between the stocks and/or areas (e.g. North Sea and herring caught in IIIa, VIaN herring and the North Sea, Celtic Sea and Irish Sea herring). Due to time constraints, not all the stock annexes were updated.

HAWG has adopted the ICES recommended procedure of benchmark and update assessments. In 2006 HAWG carried out one benchmark assessment: Celtic Sea herring. North Sea herring, VIaN herring, western Baltic spring spawning herring and North Sea sprat were all update assessments in 2007. VIaS and Irish Sea herring were all exploratory assessments. No exploration of IIIa herring was carried out in 2007. Two stocks, with very poor data (no catch at age sampling) and no current ongoing research are described in chapter 10. These are Clyde herring and sprat in the English Channel.

### 1.12 Recommendations

Please see Annex 2.

Table 1.5.1: Available disaggregated data for the HAWG per March 2007. X: Multiple spreadsheets (usually .xls); W: WG-data national input spreadsheets (xls); D: Disfad inputs and Alloc-outputs (ascii/txt)

| Stock | Catchyear |  | Comments |
| :---: | :---: | :---: | :---: |
|  |  | x W D |  |
| Baltic Sea: Illa and SD 22-24 |  |  |  |
| her_3a22 | 1991-2000 | $x$ | raw data, provided by Jorgen Dalskov, Mar. 2001, spliting revised |
|  | 1998 | x | provided by Jorgen Dalskov, Mar. 2001, spiititing revised |
|  | 1999 | x | provided by Jorgen Dalskov, Mar. 2001, splititing revised, catch data revised |
|  | 2000 | x | provided by Jorgen Dalskov, Mar. 2001 |
|  | 2001 | x | provided by Jorgen Dalskov, Mar. 2002 |
|  | 2002 | $\times$ | provided by Jorgen Dalskov, Mar. 2003 |
|  | ${ }^{2003}$ | $\times$ | provided by Jorgen Dalskov, Mar. 2004 |
|  | 2004 | $\times$ | provided by Lotte Worsse Clausen, Mar. 2005 |
|  | 2005 2006 | $\times$ | provided by Lotte Worsøe Clausen, Mar. 2006 provided by Mikael van Deurs, Mar. 2007 |
| Celtic Sea and VIIJ her_irls |  |  |  |
|  | 1999 | x | provided by Ciarán Kelly, Mar. 2000 |
|  | 2000 | $\times$ | provided by Ciarán Kelly, Mar. 2001 |
|  | 2001 | D | provided by Ciarán Kelly, Mar. 2002 |
|  | 2002 | D | provided by Ciarán Kelly, Mar. 2003 |
|  | 2003 | D | provided by Maurice Clarke, Mar. 2004 |
|  | 2004 |  | provided by Maurice Clarke, Mar. 2005 |
|  | 2005 2006 |  | provided by Maurice Clarke, Mar. 2006 provided by Maurice Clarke, Mar. 2007 |
| Clyde |  |  |  |
| her_clyd | $\begin{gathered} 1999 \\ 2000-2003 \end{gathered}$ | $\times$ | provided by Mark Dickey-Collas, Mar. 2000 included in VlaN |
| Irish Sea |  |  |  |
| her_nirs | 1988-2003 | $\times$ | updated by SG HICS, March 2004 |
|  | 1998 | $x$ | provided by Mark Dickey-Collas, Mar. 2000 |
|  | 1999 |  | provided by Mark Dickey-Collas, Mar. 2000 |
|  | 2000 | x w | provided by Mark Dickey-Collas, Mar. 2001 |
|  | 2001 |  | provided by Mark Dickey-Collas, Mar. 2002 |
|  | 2002 | $\times$ | provided by Richard Nash, Mar. 2003 |
|  | 2003 | $\times$ | provided by Richard Nash, Mar. 2004 |
|  | 2004 | $\times$ | provided by Beatriz Roel, Mar. 2005 |
|  | 2005 | $\times$ | provided by Steven Beggs, Mar. 2005 |
|  | 2006 | x | provided by Steven Beggs, Mar. 2006 |
| North Sea |  |  |  |
| her_47d3, her_nsea | 1991 | $\times$ | provided by Yves Verin, Feb. 2001 |
|  | 1992 |  | provided by Yues Verin, Feb. 2001 |
|  | 1993 | x | provided by Yves Verin, Feb. 2001 |
|  | 1994 | $\times$ | provided by Yves Verin, Feb. 2001 |
|  | 1995 | $\times \mathrm{w}$ D | provided by Yves Verin, Feb. 2001, updated by SG Rednose, Oct 2003 |
|  | 1996 | (x) w D | provided by Yees Verin, Feb. 2001, updated by SG Rednose, Oct 2003 |
|  | 1997 | (x) ${ }^{\text {w }}$ w ${ }^{\text {w }}$ | provided by Yves Verin. Feb. 2001, updated by SG Rednose, Oct 2003 |
|  | 1998 | (x) ${ }^{W}$ | provided by Yves Verin, Mar. 2000, updated by SG Rednose, OCt 2003 |
|  | 1999 | $W$ $W$ $W$ | provided by Christopher Z Zimmermann, Mar. 2000 , updated by SG Rednose, Oct 2003 provided by Christopher $\mathrm{Zimmermann}, \mathrm{Mar}. \mathrm{2001} ,\mathrm{updated} \mathrm{by} \mathrm{SG} \mathrm{Rednose}$,Oct 2003 |
|  | 2000 | $\begin{array}{ll}W \\ W \\ W & \text { D }\end{array}$ | provided by Chistopher Zimmermann, Mar. 2001 , updated by SG Rednose, Oct 2003 provided by Christopher zimmermann, Mar. 2002 |
|  | 2002 | w D | provided by Christopher Zimmermann, Mar. 2003 |
|  | 2003 | w | provided by Christopher Zimmermann, Mar. 2004 |
|  | 2004 | w D | provided by Christopher Zimmermann, Mar. 2005 |
|  | 2005 2006 | $\begin{array}{ll}W & \text { D } \\ W\end{array}$ | provided by Christopher Zimmermann, Mar. 2006 |
| West of Scotland (VIa(N)) |  |  |  |
| her_vian | 1957-1972 | x | provided by John Simmonds, Mar. 2004 |
|  | 1997 | $\times$ | providea by ken Paterson, Mar. 2002 |
|  | 1998 | $\times$ | provided by Ken Patterson, Mar. 2002 |
|  | 1999 | w D | provided by Paul Fernandes, Mar. 2000 , W included in North Sea |
|  | 2000 | W D | provided by Emma Hattield, Mar. 2001, W included in North Sea |
|  | 2001 | W <br> W <br> D | provided by Emma Hatifild, Mar. 2002, W included in North Sea provided by Emma Hatield, Mar. 203, w included in North Sea |
|  | ${ }_{2003}^{2002}$ | ${ }_{W}^{W}$ | provided by Emma Hatifild, Mar. 2003, W included in North Sea provided by Emma Hatield, Mar. 2004 , W included in North Sea |
|  | 2004 | w D | provided by John Simmonds, Mar. 2005, W included in North Sea |
|  | 2005 | w D | provided by Emma Hatield, Mar. 2006, W included in North Sea |
|  | 2006 | w D | provided by Emma Hatfield, Mar. 2007, W included in North Sea |
| $\begin{array}{c}\text { West of Ireland } \\ \text { her irlw }\end{array}$ $1999 \quad \mathrm{X}$ (W) provided by Ciaran Kelly, Mar. 2000 |  |  |  |
|  |  |  |  |  |
|  | 2000 2001 | $\times$ (w) | provided by Ciaran Kelly, Mar. 2001 |
|  | 2001 |  | provided by Ciaran Kelly, Mar. 2002 provided by Ciaran Kelly, Mar. 2003 |
|  | 2003 |  |  |
|  | 2004 | D | provided by Maurice Clarke, Mar. 2005 |
|  | 2005 |  | provided by Afra Egan, Mar. 2006 |
|  | 2006 |  | provided by Afra Egan, Mar. 2007 |
| Sprat in Illa |  |  |  |
| spr_kask |  |  | provided by Else Torstensen, Mar. 2000 |
|  | 2000 | $\times$ (w) | provided by Else Torstensen, Mar. 2001 |
|  | 2001 | $\times$ ( ${ }^{\text {( })}$ | provided by Lotte Askgaard Worsge, Mar. 2002 |
|  | 2002 2003 | $\times(w)$ $\times(w)$ | provided by Lotte Worsse Clausen, Mar. 2003 provided by Lote Worsge Clausen, Mar. 2004 |
|  | 2004 | $\times$ (w) | provided by Lotte Worsse Clausen, Mar. 2005 |
|  | 2005 | $\times$ (w) | provided by Lotte Worsse Clausen, Mar. 2006 |
|  | 2006 | $x$ (W) | provided by Mikael van Deurs, Mar. 2007 |
| Sprat in the North Sea |  |  |  |
|  | 1999 | $\times(\mathbb{W})$ $\times(W)$ | provided by Else Torstensen, Mar. 2000 |
|  | 2001 | $\times$ (w) | provided by Lotte Askgaard Worsse, Mar. 2002 |
|  | 2002 | $\times$ (w) | provided by Lotte Worsse Clausen, Mar. 2003 |
|  | 2003 | $x$ ( ${ }^{\text {a }}$ ) | provided by Lotte Worsge Clausen, Mar. 2004 |
|  | 2004 | $\times$ (w) | provided by Lotte Worsse Clausen, Mar. 2005 |
|  | 2005 | $\times$ (w) | provided by Lotte Worsse Clausen, Mar. 2006 |
| Sprat in VIld \& e <br> spr ech 1999 (W) provided by Else Torstensen, Mar. 2000 |  |  |  |
|  |  |  |  |  |
|  | 1999 | $\times$ x (W) | provided by Else Torstensen, Mar. 2000 provided by Else Torstensen, Mar. 2001 |
|  | 2001 | $\times$ (w) | provided by Lotte Askgaard Worsse, Mar. 2002 |
|  | 2002 | $x$ (w) | provided by Lotte Worsse Clausen, Mar. 2003 |
|  | 2003 | $\times$ ( ${ }^{\text {(w) }}$ | provided by Lotte Worsge Clausen, Mar. 2004 |
|  | 2004 | $\times$ (w) | provided by Lotte Worsse Clausen, Mar. 2005 |
|  | 2005 | $\times$ (w) | provided by Lotte Worsse Clausen, Mar. 2006 |
| National Data |  |  |  |
|  |  | $x$ |  |
| Germany: North Sea | ${ }_{1} 1995-1998$ | w | provided by Christopher Zimmermann, Mar 2001 (without sampling) |
| Norway: Sprat | 1995-1998 | w | provided by Else Torstensen, Mar 2001 (without sampling) |
| Sweden | 1990-2000 | w | provided by Johan Modin, Mar 2001 (without sampling) |
| UK/England \& Wales UK/Scotland | 1985-2000 | w | database output provided by Marinelle Basson, Mar. 2001 (without sampling) |

Table 1.7.1: Sampling of the pelagic fleet by country, quarter and area for the North Sea (area IV) and area VIId. No. trip $=$ number of trips. Total hauls $=$ total number of hauls sampled. Herring hauls $=$ total number of hauls sampled with herring catches (landings and/or discards) on a discard observer trip.

| 2006 | Country | QUARTER | AREA | No. TRIPS | Total hauls | HERRING HAULS |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| Scotland | 1 | IVa | 6 | 19 | 6 |  |
|  | Scotland | 2 | IVa | 3 | 4 | 4 |
|  | Netherlands | 3 | Iva,IVb |  | 38 |  |
| Netherlands | 3 | IVa |  | 54 |  |  |
| Scotland | 3 | IVa | 11 | 26 | 26 |  |
| Scotland | 4 | IVb | 13 | 39 | 3 |  |
| Netherlands | 4 | VIId | - | 33 |  |  |

this table is based on the information available at the HAWG. It should not be regarded as a complete list of all biological samples taken in the pelagic fleet. The samples taken by The Netherlands are obtained from 11 trips.

Table 1.7.2 Sampling of the pelagic fleet by country, quarter and area for the remaining areas covered by the national sampling programmes within HAWG. No. trip = number of trips. Total hauls $=$ total number of hauls sampled. Herring hauls $=$ total number of hauls sampled with herring catches (landings and/or discards) on a discard observer trip.

| 2006 | Country | Quarter | Area | No. TRIPS | Total hauls | Herring hauls |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Scotland | 1 | VIa | 7 | 19 | 2 |
|  | Scotland | 3 | VIa | 3 | 8 | 8 |
|  | Germany | 1 | VIaN | 1 | 1 | 1 |
|  | Netherlands | 1 | VIIj, VIIc |  | 28 |  |
|  | Netherlands | 1 | VIIj, VIIc, VIIb |  | 32 |  |
|  | Netherlands | 1 | VIIb, VIIh, <br> VIIj, VIIIa |  | 33 |  |
|  | Netherlands | 2 | Vla |  | 35 |  |
|  | Netherlands | 2 | VIIIa |  | 6 |  |
|  | Netherlands | 2 | VIa, IVa, Vbl |  | 53 |  |
|  | Netherlands | 3 | Iva, Via, VIIb,VIIe,VIIj |  | 39 |  |
|  | Netherlands | 4 | VIId, VIIe, VIIh |  | 40 |  |

* this table is based on the information available at the HAWG. It should not be regarded as a complete list of all biological samples taken in the pelagic fleet. The samples taken by The Netherlands are obtained from 11 trips.

Table 1.5.3.1: General sampling characteristics of the Dutch port sampling program over the years 2004, 2005 and 2006, all quarters and divisions combined.

|  | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | :---: | :---: | :---: |
| Number of samples | 163 | 136 | 136 |
| Number of fish aged | 4050 | 3400 | 3400 |
| Median age (IQR) | $5(4-6)$ | $5(5-7)$ | $4(4-6)$ |
| Average weight (SD) <br> (grams) | $187(77)$ | $213(79)$ | $190(64)$ |
| Landing weight <br> (tonnes) | 159038 | 152488 | 110404 |

Table 1.5.3.2: Results of the bootstrap algorithm to estimate precision in numbers-at-age as applied to CS herring landings in Ireland in 2006.

CS Herring 2006 Q1 (analysis id 36)

| age | mean num |  |  | min num |  |
| ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 max num | cv num |  |  |  |
|  | 7099566 | 2633281 | 15020064 | 27.35 |  |
|  | 14430365 | 11276827 | 18117289 | 7.16 |  |
|  | 3 | 1431651 | 1122161 | 1657861 | 6.22 |
|  | 4 | 8929221 | 7082874 | 10602603 | 7.14 |
|  | 5 | 3426586 | 2117355 | 4527563 | 10.5 |
|  | 6 | 1161752 | 565975 | 1602919 | 14.71 |
|  | 7 | 186414 | 66192 | 308662 | 20.33 |

CS Herring 2006 Q3 (analysis id 52)
$\mathbf{a g}$

|  | m | min num | max num | cv |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 253766 | 120043 | 356106 | 14.42 |
| 2 | 9713009 | 5162143 | 12682446 | 11.38 |
| 3 | 4640709 | 3513891 | 5192945 | 5.35 |
| 4 | 1646663 | 1372748 | 1926113 | 5.32 |
| 5 | 3570583 | 2672145 | 4460396 | 7.56 |
| 6 | 1433178 | 939268 | 2352245 | 13.92 |
| 7 | 582711 | 294003 | 1481647 | 31.54 |
| 8 | 105343 | 22056 | 411620 | 60.45 |
| 9 | 43557 | 7993 | 157754 | 55.55 |
| 1 | 9238 | 1651 | 33908 | 52.8 |

CS Herring 2006 Q4 (analysis id 32)

| age | mean num | min num | max num | cv num |
| :---: | :---: | :---: | :---: | :---: |
| 0 | 13561 | 7113 | 52135 | 51.74 |
| 1 | 829905 | 633386 | 1038518 | 7.05 |
| 2 | 11601521 | 9088740 | 13499252 | 6.05 |
| 3 | 3004161 | 2500185 | 3341010 | 4.39 |
| 4 | 655337 | 410789 | 1005532 | 14.85 |
| 5 | 1548625 | 1037762 | 2230587 | 12.75 |
| 6 | 614223 | 319570 | 1062886 | 20.45 |
| 7 | 196611 | 87246 | 376347 | 25.53 |
| 8 | 27034 | 14942 | 52978 | 24.76 |
| 9 | 8099 | 572 | 25594 | 57.95 |

Table 1.5.3.3: Results of the bootstrap algorithm to estimate precision in numbers-at-age as applied to NW herring landings in Ireland in 2006.

## NW Herring 2006 Q1 (analysis id 50)

| age | mean num | min num | max num | cv num |
| :---: | :---: | :---: | :---: | :---: |
| 2 | 905576 | 478885 | 1458279 | 17.8 |
| 3 | 20332250 | 12820605 | 29427317 | 10.66 |
| 4 | 19160703 | 18134865 | 20227194 | 1.68 |
| 5 | 14591315 | 13041387 | 16136608 | 2.92 |
| 6 | 6396043 | 4186431 | 7979064 | 7.61 |
| 7 | 1718249 | 778715 | 2380277 | 12.36 |
| 8 | 597265 | 179377 | 958553 | 17.75 |
| 9 | 190492 | 50673 | 363675 | 22.28 |
| 10 | 79568 | 16130 | 203110 | 37.7 |

NW Herring 2006 Q4 (analysis id 49)

| aque | mean num | min num | max num | cv num |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 599724 | 144769 | 1366649 | 32.06 |
| 2 | 21329275 | 13423887 | 30397002 | 13.61 |
| 3 | 17330775 | 13808288 | 19742328 | 4.29 |
| 4 | 10556765 | 8390400 | 12088730 | 5.66 |
| 5 | 7635251 | 5581843 | 10284265 | 10.34 |
| 6 | 2293685 | 1337493 | 3631527 | 16.07 |
| 7 | 668239 | 357452 | 1156314 | 18.77 |
| 8 | 175821 | 63514 | 390571 | 292 |
| 9 | 37275 | 6336 | 95029 | 37.45 |

Table 1.8.3.1 : Time series used in the analysis.

| Stock | Length of the time series | Age of recruits | Source |
| :--- | :---: | :---: | :---: |
| North Sea | $1947-2005$ | 1 | ICES 2006 |
| Irish Sa | $1962-2005$ | 1 | ICES 2006 |
| Celtic Sea | $1959-2005$ | 1 | ICES 2006 |
| VIa(N) | $1959-2005$ | 1 | ICES 2006 |
| VIa(S) | $1971-2004$ | 1 | ICES 2006 |

Table 1.8.3.2: Parameters for the hockey stick and the Ricker stock recruitment relationship used in the analysis.

|  | Hockey stick |  | Ricker |  |
| :--- | :---: | :---: | :---: | :---: |
| Stock | $\alpha$ | $\beta$ | $\alpha$ | $\beta$ |
| North Sea | 49.10506 | 0.879536 | 63.43208 | 0.49090 |
| Irish Sa | 34.28372 | 0.005509 | 15.81970 | 18.46749 |
| Celtic Sea | 7.36974 | 0.058526 | 10.25259 | 8.26308 |
| VIa(N) | 4.84980 | 0.622325 | 5.53721 | 0.49954 |
| VIa(S) | 6.22671 | 0.119607 | 9.33552 | 4.46681 |

ICES-Rectangle


Figure 1.5.1: ICES areas as used for the assessment of herring stocks south of $62^{\circ} \mathbf{N}$. Area names in italics indicate the area separation applied to the commercial catch and sampling data kept in long term storage. "Transfer area" refers to the transfer of Western Baltic Spring Spawners caught in the North Sea to the Baltic Assessment.


Fig 1.5.3.1. Observed CV of weight-at-age in relation to the number of fish sampled per age class as derived from the Swedish port sampling program in 2006.


Fig 1.5.3.2. Schematic outline of the Dutch port sampling scheme for commercial herring landings.


Figure 1.5.3.3. Estimates of numbers-at-age together with their (relative) standard error as derived from the Dutch port sampling program in 2004.


Figure 1.5.3.4. Estimates of numbers-at-age together with their (relative) standard error as derived from the Dutch port sampling program in 2005.


Figure 1.5.3.5. Estimates of numbers-at-age together with their (relative) standard error as derived from the Dutch port sampling program in 2006.


Figure 1.8.1.1. North Sea IBTS catches of anchovy and sardine.


Figure 1.8.1.2. North Sea herring stock to recruit relationship (with Ricker curve) for the time series 1960-2006 from the standard assessment using a annual fixed $M$ that varies by age and an annually varying $M$ from the ICES multispecies working group.


Figure 1.8.3.1: The net and surplus production of the original data output of ICA final run 2006 and the hockey stick and the Ricker stock recruitment relationship adjusted to calculate the residuals of the recruits of each year used as a weighing factor for the calculation of the net and surplus production assuming different $F$; The recruit per spawner ratio is shown for comparison.


Figure 1.8.3.2: Recruit per spawner from the original ICA output 2006 and the hockey stick and the Ricker stock recruitment relationship adjusted to calculate the residuals of the recruits of each year used as a weighing factor for the calculation of the net and surplus production assuming a $F$ of 0.25 for the Irish Sea, the Celtic Sea, the VIa(N) and the VIa(S) herring stock.


Figure 1.10.1 WG estimates of catch (yield) of the stocks presented in HAWG 2007.


Figure 1.10.2: Spawning stock biomass estimates of the 3 stocks for which assessments were presented in HAWG 2007. The $B_{p a}$ level (if defined) is indicated in the graphs.


Figure 1.10.3 Estimates of mean $F$ of the 3 stocks for which analytical assessments were presented in HAWG 2007. The $F_{p a}$ level (if defined) is indicated in the graphs.

### 2.1 The Fishery

### 2.1.1 ACFM advice and management applicable to 2006 and 2007

According to the management scheme agreed between the EU and Norway, adopted in December 1997 and last amended in November 2004, efforts should be made to maintain the SSB of North Sea Autumn Spawning herring above 800000 tonnes. An SSB reference point of 1.3 million has been set $\left(=\mathbf{B}_{\mathrm{pa}}\right)$ above which the TACs will be based on an $\mathrm{F}=0.25$ for adult herring and $\mathrm{F}=0.12$ for juveniles. If the SSB falls below 1.3 million tonnes, the fishing mortality will have to be linearly reduced. A TAC deviation of more than $15 \%$ between two subsequent years should be avoided, however, the TAC might be reduced by more than $15 \%$ if the parties consider this appropriate.

Since 2002, the SSB is considered to have been above $\mathbf{B}_{\mathrm{pa}}$. From then on, ACFM gave fleetwise catch option tables for fishing mortalities within the constraints of the EU-Norway management scheme. The advice for a sub-TAC on catches in IVc and VIId for 2004 was that it should not increase faster than the TAC for the North Sea as a whole. ACFM thought that a share of $11 \%$ on the total North Sea TAC (average share 1989-2002) would be an appropriate guide to distributing the harvesting of Downs herring.

It was expected at that time that fishing at the recommended level would lead to a further increase in the SSB in the short term, mainly due to large recruiting year classes entering the fishery. ACFM considered in 2006 that there were four recruiting year classes (2002, 2003, 2004 and 2005) that were all well below average. Last year ACFM offers options of $15 \%$ to 25 \% varying TACs to managers, taking into account an increased risk that the stock may fall below the 1.3 mill. tonnes in the medium-term if the rule of $15 \%$ constraint on TAC variation is applied.

The final TAC adopted by the management bodies for 2006 was 454800 t for Area IV and Division VIId, whereof not more than 50000 t should be caught in Divisions IVc and VIId. For 2007, the TAC was reduced by $25 \%$ to 341100 t ( 37517 t in Divisions IVc and VIId).

Catches of herring in the Thames estuary are not included in the TAC. The by-catch ceiling set for fleet B in the North Sea was 42500 t for 2006 and was decreased by $25 \%$ to 31900 t for 2007. As North Sea autumn spawners are also caught in Division IIIa, regulations for the fleets operating in this area have to be taken into account for the management of the WBSS stock (see Section 3). For a definition of the different fleets harvesting North Sea herring see the stock annex and Section 2.7.2.

Following the apparent recovery of the autumn spawning North Sea herring, some regulatory measures were amended in 2004: The total Norwegian quota and half of the EU quota for Division IIIa could be taken in the North Sea. A licence scheme introduced in 1997 by UK/Scotland to reduce misreporting between the North Sea and VIaN was relaxed. The minimal amount of target species in the EU industrial fisheries in IIIa has been reduced to 50 \% (for sprat, blue whiting and Norway pout). Since 2005, for Division IIIa, Norway could only take half of its quota in the North Sea, and there is no flexibility for EU vessels. These amendments were kept for 2006. For 2007, Norway could take $40 \%$ of the IIIa quota in the North Sea.

### 2.1.2 Catches in 2006

Total landings and estimated catches are given in the Table 2.1.1 for the North Sea and for each Division in Tables 2.1.2 to 2.1.5. Total working group catches per statistical rectangle and quarter are shown in Figures 2.1.1 a - d, the total for the year in Figure 2.1.1e. Each nation provided most of their catch data (either official landings or working group catch) by statistical rectangle.

The catch figures in Tables 2.1.1-2.1.5 are mostly provided by WG members and may or may not reflect national catch statistics. These figures can therefore not be used for legal purposes. For corrections applied to and inconsistencies in previous year's data see Section 2.2.3. Denmark and Norway provided information on by-catches of herring in the industrial fishery. These are taken in the small-meshed fishery (B-fleet) under an EU quota by Denmark and are included in the A-fleet figures for Norway. Catch estimates of herring taken as bycatch by other small-mesh fisheries in the North Sea may be an underestimate. The total catch in 2006 as used by the Working Group amounted to 510600 t .

Total herring catches (including industrial catches as well) by area are stable in the most recent year in Division IVa (East) (+ 3 \%) and IVb (almost no change). Catches decreased in IVa (West) by $31 \%$ (after an increase of $40 \%$ in 2005) and by $24 \%$ in the southern North Sea (Division IVc and VIId).

Landings of herring taken as by-catch in the Danish small-meshed fishery in the North Sea have decreased by $45 \%$ to 11900 t as compared to last year (Table 2.1.6). These industrial herring catches were much lower than the by-catch ceiling set by the EU (42 500 t ). In 2006, the Danish sprat fishery was carried out throughout the year with by-catches of herring of about 7 \% (8 983 t ; by-catch 2005: 9\%). In the Norwegian industrial fishery, herring by-catch is almost the same in $2006(961 \mathrm{t})$, compared to 998 t last year. The relative small proportion of herring by-catch could be influenced by the closure of the Norway pout fishery. The quarterly distribution of herring by-catches in the Norwegian industrial fishery and its relative share on the total industrial landings are given in the text table below. These figures are counted against the human consumption quota.

| QUARTER 1 | QUARTER 2 | QUARTER 3 | QUARTER 4 | TotaL |
| :---: | :---: | :---: | :---: | :---: |
| 49 t | 314 t | 27 t | 571 t | 961 t |
| $2.6 \%$ | $0.7 \%$ | $0.1 \%$ | $2.9 \%$ | $0.8 \%$ |

There is not much information on misreporting of landings taken in the North Sea but reported from other areas available. Misreporting within the North Sea accounts to 18800 t .

Based on WG estimates of total catch, TACs for the human consumption fishery in Subarea IV and Division VIId have been greatly exceeded for several years. This appears to have continued in 2006, but on a somewhat lower level. Catches in the human consumption fishery have reduced to 498000 t (decreased by $19 \%$ ) in 2006, so the excess over the TAC for the human consumption fishery amounted to 43000 t ( 9 \%) in the most recent year.

The total catch in the North Sea was 510600 t , the TAC was set to 497300 t (HC and industrial fishery). The over catch of total TAC in 2006 amounted to 13300 t (<3\%). While the TAC in southern North Sea and the Eastern Channel was met well in 2004 and 2005, there is an over catch in 2006 of 6600 t in this area ( $13 \%$ ). So misreporting is likely to be spread equally between the areas in the North Sea.

The total North Sea TAC excess for the years 1995 to 2006 is shown in the table below (adapted from Table 2.1.6). Since the introduction of yearly by-catch ceilings in 1996, these ceilings have never been exceeded.

| YeAR | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TAC HC ('000 t) | 156 | 159 | 254 | 265 | 265 | 265 | 265 | 400 | 460 | 535 | 455 |
| "Official" landings HC ('000 t) ${ }^{1}$ | 170 | 162 | 253 | 275 | 267 | 275 | 282 | 414 | 484 | 547 | 478 |
| Working Group catch HC ('000 t) | 196 | 226 | 324 | 318 | 328 | 303 | 331 | 438 | 537 | 617 | 498 |
| Excess of landings over TAC HC ('000 t) | 40 | 67 | 70 | 53 | 63 | 38 | 66 | 38 | 77 | 83 | 43 |
| By-catch ceiling ('000 t) ${ }^{3}$ | 44 | 24 | 22 | 30 | 36 | 36 | 36 | 52 | 38 | 50 | 42 |
| Reported by-catches ('000 t) ${ }^{4}$ | 38 | 13 | 14 | 15 | 18 | 20 | 22 | 12 | 14 | 22 | 12 |
| Working Group catch North Sea ('000 t) | 233 | 238 | 338 | 333 | 346 | 323 | 353 | 450 | 550 | 639 | 511 |

HC = human consumption fishery
1 "Official" landings might be provided by WG members; they do not in all cases correspond to official catches and cannot be used for management purposes. Norwegian by-catches included in this figure.
${ }^{2}$ figure altered in 2000 on the basis of a re-evaluation of misreported catches from VIa North.
${ }^{3}$ by-catch ceiling for EU industrial fleets only, Norwegian by-catches included in the HC figure.
${ }^{4}$ provided by Denmark only.

### 2.2 Biological composition of the catch

Biological information (numbers, weight, catch (SOP) at age and relative age composition) on the catch as obtained by sampling of commercial catches is given in Tables 2.2.1 to 2.2.5. Data are given for the whole year and by quarter. Except in cases where the necessary data are missing, data are displayed separately by area for herring caught in the North Sea, Western Baltic spring spawners (only in IVaE), and the total NSAS stock, including catches in Division IIIa.

Biological information on the NSAS caught in Division IIIa was obtained using splitting procedures described in Sec. 3.2 and in the stock annex 2. Note that splitting was only applied to the working group catch, following the correction of area misreporting.

The Tables are laid out as follows:
Table 2.2.6: Total catches of NSAS (SOP figures), mean weights and numbers-at-age by fleet

Table 2.2.7: Data on catch numbers-at-age and SOP catches for the period 1991-2006 (herring caught in the North Sea)

Table 2.2.8: WBSS taken in the North Sea (see below)
Table 2.2.9: NSAS caught in Division IIIa
Table 2.2.10: Total numbers of NSAS
Table 2.2.11: Mean weights-at-age, separately for the different Divisions where NSAS are caught, for the period 1996-2006.

Note that SOP catch estimates may deviate in some instances slightly from the working group catch used for the assessment.

### 2.2.1 Catch in numbers-at-age

The total number of herring taken in the North Sea and the total number of NSAS have decreased by 30 \% (to 3.7 billion fish) and by 25 \% (to 3.8 billion fish), respectively, as compared to last year. 0 - and 1-ringers contributed $25 \%$ of the total catch in numbers of NSAS in 2006 (Table 2.2.7). 0- and 1-ringer catch has decreased by $10 \%$ and almost $80 \%$, respectively, as compared to 2005. Figure 2.2.1. shows the relative proportions of the total catch numbers for different periods. The catches contain more than $50 \%$ of the age group $4+$ winter ringers. This is consistent in all area in the North Sea. Catches are still dominated by
the 2000 year class, but the catches also show larger quantities of 0 ringers. In area IVc and VIId, 0 and 1-winter ringers accumulate only to less than $10 \%$ in the catch.

The following table summarises the total catch in tonnes of North Sea autumn spawners. To arrive at the total catch of NSAS, splitting of the catch into NSAS and Western Baltic Spring Spawners has to be done in Divisions IIIa and IVaE. WBSS from the North Sea are then subtracted and NSAS from IIIa added to the total NSAS catch figure. The final total catch used for the assessment of NSAS in 2006 was 515000 tonnes:

| Area | Allocated | Unallocated | Discards | Total |
| :---: | :---: | :---: | :---: | :---: |
| IVa West | 243559 | 10981 | 1492 | 256032 |
| IVa East | 102628 | - | - | 102628 |
| IVb | 92996 | 2364 | - | 95360 |
| IVc/VIId | 51178 | 5419 | - | 56597 |
|  | Total catch in the | Sea |  | 510617 |
|  | Autumn Spawners caught in Division IIIa (SOP) |  |  | 15015 |
|  | Baltic Spring Spawners caught in the North Sea (SOP) |  |  | -10 953 |
|  | Other Spring Spawners |  |  | -65 |
|  | Total Catch NSAS used for the assessment |  |  | 514614 |

"Other spring spawners" are 65 t of Blackwater herring caught under a separate quota and included in the catch figure for England \& Wales. This year no spring spawners were reported from the commercial catch taken in other areas of the Western North Sea (see Sect. 2.2. below).

### 2.2.2 Spring-spawning herring in the North Sea

Norwegian Spring-spawners and local fjord-type spring spawning herring are taken in Division IVa (East) close to the Norwegian coast under a separate TAC. These catches are not included in the Norwegian North Sea catch figures given in Tables 2.1.1 to 2.1.6, but are listed separately in the respective catch tables. The amount of these catches varied significantly between less than 626 t in 2005 and 55000 t in 1997. Coastal Spring Spawners in the southern North Sea (e.g. Thames estuary) are caught in small quantities (usually less than 100 t ) regulated by a local TAC. The Netherlands reported increasing catches of spring spawners in the Western Part of the North Sea in some years, which were included in the national catch figures and subtracted from the total catch used for the assessment of NSAS, but in the last three years no spring spawners were reported from routine sampling of commercial catch taken in the west.

Western Baltic and local Division IIIa Spring-spawners (WBSS) are taken in the eastern North Sea during the summer feeding migration (see stock annex 2 and section 3.2.2). These catches are included in Table 2.1.1 and listed as IIIa type. Table 2.2.8 specifies the estimated catch numbers of WBSS caught in the North Sea, which are transferred from the North Sea assessment to the assessment of Division IIIa/Western Baltic in 1991-2006.

The method of separating these fish, using vertebral counts as described in former reports of this Working Group (ICES 1991/ACFM:15), is given in detail in Section 3.2.2.1 and in stock annex 2. The source for the splitting were samples taken from Danish and Norwegian catches, obtained in all quarters. The mean vertebral counts for herring 2 -ringers, 3 -ringers, and 4+ringers caught in the $2^{\text {nd }}$ and $3^{\text {rd }}$ quarter in the transfer area are given in Figure 2.2.2. Details for the splitting procedure are given in section 3.2.2.1 The resulting proportion of spring spawners and the quarterly catches of these in the transfer area in 2006 were as follows:

| QUARTER | 1- <br> RINGERS <br> $(\%)$ | 2- <br> RINGERS <br> $(\%)$ | 3- <br> RINGERS <br> $\mathbf{( \% )}$ | 4+- <br> RINGERS <br> $(\%)$ | CATCH In THE <br> TRANSFER AREA (t) | CATCH OF WBSS in THE <br> NORTH SEA (t) |
| :--- | :---: | ---: | :---: | ---: | ---: | ---: |
| Q 1 | $50 \%$ | $23 \%$ | $25 \%$ | $2 \%$ | 4951 | 249 |
| Q 2 | $28 \%$ | $36 \%$ | $53 \%$ | $57 \%$ | 13055 | 7214 |
| Q 3 | $14 \%$ | $3 \%$ | $74 \%$ | $62 \%$ | 3687 | 2203 |
| Q 4 | $3 \%$ | $74 \%$ | $62 \%$ | $26 \%$ | 4857 | 1288 |
| total |  |  |  |  | 26551 | 10953 |

The quarterly age distribution and mean weight-at-age in sub-division IVa East was applied to the catches of the first, second, third, and fourth quarter in the transfer area. The numbers of spring spawners by age were obtained by applying the estimated proportion by age.

### 2.2.3 Data revisions

The result of the splitting procedure in 2006 for the transfer area is also known to contain a small bug in terms of tonnage. This was estimate to be below $0.5 \%$, therefore the table was not updated after this failure was recognized. This will be done in next years assessment.

There were two revisions to the historic catch data time series carried out this year. An incorrect allocation of fish to the plus group in the Dutch catches in 2004 and 2005 affected the age distribution in both years. In 2006, new sources of information on catch misreporting from the UK became available. This information was associated with a stricter enforcement regime that may be responsible for the lack of area misreporting between area IV and VIa(N) in 2006 (see also section 5.1.3). In light of this new information on misreporting a readjustment of catch figures was necessary from 2001 to 2004. The resulting changes to the catch in numbers, catch in tonnes and the mean weights at age in the catch are documented in Table 2.2.13.

### 2.2.4 Quality of catch and biological data, discards

As in previous years, some nations provided information on misreported and unallocated catches of herring in the North Sea and adjacent areas. Catches made in Division IVa were mainly misreported to Division VIa, IIIa and IIa, but misreporting also occurred from IIIa to IVa, within Area IV, and from Division VIId to IVb. The Working Group catch, which includes estimates of discards and misreported or unallocated catches (see Section 1.5), was estimated to exceed the official catch by $3 \%$. It is likely that this figure is an underestimate as it only includes information from a fraction of the fleets fishing herring in the North Sea, as an analysis conducted in 2002 indicated (ICES 2002/ACFM:12). This corroborates suggestions of the Study Group for Herring Assessment Procedures (ICES 2001/ACFM:22), that a important uncertainty of the total catch figure exists since the re-opening of the fishery in 1980.

Information on discards is rare in 2006. The final figure for discards as used in the assessment was 1492 t , based on the raised discards for one fleets. As discards are likely to occur in all nation's fisheries, this figure is certainly an underestimate. Discard data has not been consistently available for the whole time series and was only included in the assessment when reported. Estimates of discards in the Dutch fleet are in the order of 5000 t per year, but cannot be split between area IV and VIaN. These are not included in the assessment.

The European Union implemented a new sampling regime in 2002, obliging member states to meet specified overall sampling levels. However, the sampling of commercial landings in 2006 for herring length and weight measured has decreased by $25 \%$ when compared to 2005, while the number of age readings has increased by 50 \% (Table 2.2.12). Only $79 \%$ of the total catch was sampled in 2006 (2005: 95 \%). It should be observed that "sampled catch" in Table 2.2.12 refers to the proportion of the reported catch to which sampling was applied. This
figure is limited to 100 \% but might in fact exceed the official landings due to sampling of discards, unallocated and misreported catches.

However, more important than a sufficient overall sampling level is an appropriate spread of sampling effort over the different metiers (each combination of fleet/nation/area and quarter). Of 107 different reported metiers, only 39 were sampled in 2006 . Some of them, however, yielded very little catch. The recommended sampling level of more than 1 sample per 1000 t catch has been met only for 19 metiers (2005:14). For age readings (recommended level $>25$ fish aged per 1000 t catch) this is also worse: only 21 metiers appear to be sampled sufficiently (2005: 17). The catch of France, UK/England and Wales, Sweden, UK/Northern Ireland, the Faroe Islands and Belgium from the North Sea has not been sampled. Information on catches landed abroad was again not available or could not be used. While it is known that by-catches of herring in other than the directed human consumption fisheries occur, most countries have not implemented a sampling scheme for monitoring these fisheries.

The WG recommends that all metiers with substantial catch should be sampled (including bycatches in the industrial fisheries), and that catches landed abroad should be sampled and information on these samples should be made available to the national laboratories (see Section 1.5).

### 2.3 Fishery Independent Information

### 2.3.1 Acoustic Surveys in VIa(N) and the North Sea in July 2006

Five surveys were carried out in the North Sea during late June and July 2006 covering most of the continental shelf north of $51^{\circ} 30^{\prime} \mathrm{N}$ in the North Sea and $56^{\circ} \mathrm{N}$ to the west of Scotland to a northern limit of $62^{\circ} \mathrm{N}$. The eastern edge of the survey area is bounded by the Norwegian, Danish, Swedish, German and Dutch coasts. The western edge is bounded by the UK coast and by the shelf edge at approximately 200 m depth. The individual surveys and the survey methods are given in the report of the Planning Group for Herring surveys (ICES 2007/LRC:04). The vessels, areas and dates of cruises are given in Table 2.3.1.1 and in Figure 2.3.1.1.

The data has been combined to provide an overall estimate of numbers-at-age, maturity ogive and mean weights-at-age are calculated as weighted means of individual survey estimates by ICES statistical rectangle. The weighting applied is proportional to the survey track for each vessel that has been covered in each statistical rectangle.

## Combined Acoustic Survey Results for the North Sea:

The estimate for North Sea autumn spawning herring is shown in Table 2.3.1.2. The estimates of SSB are reasonably consistent with previous years, at 2.1 million tonnes and 11,830 million herring (Table 2.3.1.2). The survey again shows two well-above average year classes of herring (1998 and 2000), followed by smaller year classes.

The abundance of the 2004 year class which is seen in this survey for the first time is similar in magnitude to the 2001-2003 year classes. Growth of the 2000 year class seems still to be slower than average; individuals of this year class have almost the same mean length and weight as those one year younger (2001 year class).

The spatial distribution of the abundance (numbers and biomass) of autumn spawning herring are shown in Figure 2.3.1.2. The distribution of numbers by age is shown in Figure 2.3.1.3 for 1 ring, 2 ring and $3+$ ring autumn spawning herring. The survey provides estimates of maturity and weight at age: the mean weight at age for 1 and 2 ring herring along with the proportion mature for 2 and 3 ring herring are shown in Figure 2.3.1.4. The spatial distribution of mature and immature autumn spawning herring is shown in Figures 2.3.1.5 \& 2.3.1.6 respectively.

The acoustic survey estimates for 2005 have been revised following checks in PGHERS (ICES 2007). The new values are given in Table 2.3.1.3, and the changes expressed as a percentage in Table 2.3.1.4. Changes to estimated SSB are about $+3 \%$ in total.

The time series of abundance for North Sea autumn spawners, including changes to the 2005 estimate, are given in Table 2.3.1.5.

## Reference

ICES 2007 Report of the planning group for herring surveys, ICES CM2007/LRC:04.

### 2.3.2 Larvae surveys

In 2006/07 the Netherlands and Germany carried out larvae surveys and managed to cover seven out of ten areas described in the protocol. The survey effort was comparable to previous years. The areas and time periods (including numbers of samples, vessel-days in sampling and area coverage) are given in Table 2.3.2.1 and Table 2.3.2.2. The spatial extent of the surveys is shown in Figures 2.3.2.1-2.3.2.7. The historical background of the larvae surveys and the methods used for abundance calculation are described in the handbook for quality control (Appendix 2). A more detailed description is available in the manual for the international herring larvae surveys in the North Sea (ICES CM 2005/LRC: 04).

In 2006 the Orkney/Shetland area was covered in both periods. In the first period a high number of larvae were observed west and northeast of the Orkney Islands (Figure 2.3.2.1). This is the first time that this period has been surveyed in the last 11 years. The total abundance therefore is not really comparable to previous years. Apart from 1989, the current estimate was similar to those in the late 1980's, when this period was last fully surveyed. In the second period the distribution was comparable to the years before with aggregations in the North and East of the Orkney Islands and South off the Scottish coast (Figure 2.3.2.2). Although the overall abundance showed large fluctuations during the last decades, the 2006 value followed the declining trend of the last two years and represented the lowest value since 1996 (Table 2.3.2.3).

In the Buchan area (Figure 2.3.2.3) the larvae were concentrated at only two stations resulting in a low abundance estimate at $20 \%$ of the previous year value (Table 2.3.2.3).

The abundance in the Central North Sea decreased to $50 \%$ of the previous year estimate (Figure 2.3.2.4, Table 2.3.2.3).

Abundance estimates from the three surveys in the Southern North Sea resulted in a high index value, and in contrast to the downward trend in the previous year showed an increasing trend again (Tab. 2.3.2.3). The peak of the spawning activity appeared to shift towards the end of December, although the value was mainly driven through one station. Larvae were almost exclusively found in subdivision VIId (Figures 2.3.2.5-7).

The trends in the four survey areas are very different (Figure. 2.3.2.8) with an increase in the South and decline in the Central North Sea.

The model for the Multiplicative Larval Abundance Index (MLAI) was fitted to abundances of larvae less than 10 mm in length ( 11 mm for SNS). The analysis of variance and the parameter estimates are given in Table 2.3.2.4. The updated MLAI time-series is shown in Table 2.3.2.5. The estimated trend in spawning stock biomass from this model fit is plotted in Figure 2.3.2.9 along with the SSB values obtained from the ICA runs of the Herring Assessment Working Group.

Both the LAI per unit as well as the MLAI from the larvae surveys in period 2005/2006 indicate that the SSB has decreased considerably when compared to last years WG estimate (Table 2.3.2.5).

### 2.3.3 International Bottom Trawl Survey (IBTS)

The International Bottom Trawl Survey (IBTS) started out as a young herring fish survey in 1966 with the objective of obtaining annual recruitment indices (abundance of 1-ringers in $1^{\text {st }}$ quarter) for the combined North Sea herring stock. It has been carried out every year since, and presently the survey provides recruitment indices not only for herring, but for roundfish species as well. Examinations of the catch of adult herring during the $1^{\text {st }}$ quarter IBTS have shown that this catch also indicates abundances of 2-5+ herring. During night-time on the IBTS $1^{\text {st }}$ quarter, additional sampling of herring larvae (0-ringers) is carried out by small, finemeshed nets. From 1977 to 1991 the gear was a small mid-water trawl (IKMT), but due to poor catchability of this gear, the standard gear was changed to a 2 metre ring net (MIK), used since the 1991 sampling. The total abundance of herring larvae in the survey area is used as an estimate of 0 -ringer abundance of the stock. Hence, a series of herring abundance indices (05+ ringers) are available from the IBTS programme.

### 2.3.3.1 Indices of 2-5+ ringer herring abundances

Fishing gear and survey practices were standardised from 1983, and the series of 2-5+ ringer abundance estimates from 1983 onwards has shown the most consistent results in assessments of these age groups. This series is subsequently used in North Sea herring assessment. Note that the abundances in Division IIIa are not included in these 2-5+ ringer indices. The IBTS time series of indices has been revised and Table 2.3.3.1 shows the new time-series of abundance estimates of 2-5+ ringers from the $1^{\text {st }}$ quarter IBTS for the period 1983-2007, while Table 2.3.3.2 contains area-disaggregated information on the IBTS indices for year 2007. This years indices are outstandingly low; the WG investigated this, but did not find any other reason than low abundances in the survey.

### 2.3.3.2 Index of 1-ringer recruitment

The 1-ringer index of recruitment is based on trawl catches in the entire survey area. The time series of indices has been revised and a new series are available for year classes 1977 to 2005 (Table 2.3.3.3). This year's estimate of the 2005 year class strength (1336) indicates a low recruitment, however higher than the preceding three year classes

Figure 2.3.3.1 illustrates the spatial distribution of 1-ringers as estimated by the trawling in February 2005, 2006 and 2007. In 2007 the main concentrations of 1-ringers were found in the areas of Great Fisher Bank, northern Dogger Bank, and in coastal areas of Kattegat. The concentrations in the North Sea are more offshore than observed the preceding years. The mean length of 1-ringer herring in the areas of peak abundance is in the order of 15 cm (Figure 2.3.3.2).

The Downs herring hatch later than the autumn spawned herring and generally appears as a smaller sized group during the $1^{\text {st }}$ quarter IBTS. A recruitment index of smaller sized 1-ringers is calculated based on abundance estimates of herring $<13 \mathrm{~cm}$ (see discussion of procedures in earlier reports (ICES CM 2000/ ACFM:12, and ICES CM 2001/ ACFM:12).

Table 2.3.3.3 includes abundance estimates of 1-ringer herring smaller than 13 cm , based on a standard retrieval of the IBTS database, i.e. the standard index is in this case calculated for herring $<13 \mathrm{~cm}$ only. Indices for these small 1-ringers are given either for the total area or the area excluding division IIIa, and their relative proportions are also shown. In the time-series, the proportion of 1-ringers smaller than 13 cm (of total catches) is in the order of $20 \%$, and the
contribution from division IIIa to the overall abundance of $<13 \mathrm{~cm}$ herring varies markedly during the period. (Table 2.3.3.3)

About $23 \%$ of this year's group of 1-ringers is smaller than 13 cm . These are almost exclusively found in the North Sea area (Table 2.3.3.3)

### 2.3.3.3 The MIK index of 0-ringer recruitment

This year's 0 -ringer index is based on 636 depth-integrated hauls with a 2 metre ring-net (the MIK). Index values are calculated as described in the WG report of 1996 (ICES 1996/ACFM:10). The series of estimates is shown in Table 2.3.3.4, the new index value of 0ringer abundance of the 2006 year class is estimated at 37.2.

The index is the lowest since the estimate of the 1989 year class strength, and it continues a now 5 year long series of low recruitment estimates (the average for these 5 years is about $50 \%$ of the all-year average). The 0 -ringers were predominantly distributed in two concentrations, one off the Scottish coast (in the central-western area) and one in the Southern Bight. Compared to the preceding year classes, which are also shown in Figure 2.3.3.3, the distribution of 0 -ringers from this year class is very restricted, without significant concentrations along the English coast. The long term trend in the distributional patterns of 0ringers is illustrated by the changes in absolute and relative abundance of 0-ringers in the western part of the North Sea (Figure 2.3.3.4). In the Figure 2.3.3.4 the relative abundance is given as the number of 0 -ringers in the area west of $2^{\circ} \mathrm{E}$ relative to the total number of 0ringers in the given year class. Since the year class 1982, when the relative abundance was $25 \%$, a general increase in abundance has been seen for the western part. In the last decade, the majority of 0 -ringers has been distributed in this area, and the calculated relative abundance of $86 \%$ for the present year class is in accordance with the long term trend.

### 2.4 Mean weights-at-age and maturity-at-age

### 2.4.1 Mean weights-at-age

Table 2.4.1.1 shows the historic mean weights-at-age (wr) in the North Sea stock during the 3rd quarter in Divisions IV and IIIa for the period 1996 to 2006. These values were obtained from the acoustic survey. The data for 2006 are taken from Table 2.3.1.2. In this quarter most fish are approaching their peak weights just prior to spawning. The spatial distribution of mean weight for 1 and 2-ringers is given in Figure 2.3.1.2. This spatial variability of mean weight is considerable but is not unusual. For comparison with the acoustic survey estimates, the mean weights-in-the-catch from the last ten years are also shown in Table 2.4.1.1 (from Section 2.2.1 for the 2006 values). For 4-ringers and older the mean weights for 2006 in the catch and acoustic survey are close to the long-term lows. For 5 -ring herring, the acoustic survey shows mean weights that are the lowest for the last 10 years and are similar to one year younger herring, supporting the view that the exceptional 2000 year class is growing slowly. This year class, possibly the largest in recent years and the first large one competing with an already large herring stock biomass, has grown more slowly than earlier year classes.

### 2.4.2 Maturity ogive

The percentages of North Sea autumn-spawning herring (at age) that spawned in 2006 were estimated from the July acoustic survey (Table 2.4.2.2). The values were determined from samples of herring from the research vessel catches examined for maturity stage, and raised by the local abundance. All herring at maturity stage between 3 and 6 inclusive (using an 8 -stage scale) in June or July were assumed to spawn in the autumn. The method and justification for the use of values derived from a single year's data was described fully in ICES (1996/ACFM:10). The values for 2- \& 3-ringers are taken from the acoustic survey results
which are presented in Table 2.3.1.2. For 2 and 3 ringers the proportion mature at $66 \%$ and $88 \%$ respectively is low for these age groups but not exceptional. The 2000 year class, which matured even more slowly, is now fully mature.

### 2.5 Recruitment

Information on the development in North Sea herring recruitment is available from the two IBTS indices, the 1 -ringer and the 0 -ringer index. Further, the ICA assessment provides estimates of the recruitment of herring in which information from the catch and from all fishery independent indices is incorporated.

### 2.5.1 Relationship between the MIK 0 -ringer and the IBTS 1 -ringer indices

The 0-ringer MIK index predicts the year class strength one year before the information is available from the IBTS 1-ringer estimates. The relationship between year class estimates from the two indices is illustrated in Figure 2.5.1 and described by the fitted linear regression. Last years prediction of the 2005 year class was confirmed by this year's IBTS 1-ringer index of the year class (black square in the figure). The good correlation between the indices is also evident when comparing the respective trends in indices during the period (Figure 2.5.2).

### 2.5.2 Trends in recruitment from the assessment

Recruitment is estimated in the ICA-assessment, and in Figure 2.5.3 the trends in 1-ringer recruitment based on 2007 assessment is illustrated. The recruitment declined during the sixties and the seventies, followed by a marked increase in the early eighties. After the strong 1985 year class recruitment declined again until the strong year classes 1998-2001. However, the 1-ringer recruitments of the recent 2002-2005 year classes are low, and the MIK index of 0 -ringer recruitment for the present year indicates a very small 2006 year class. The present ICA estimates of 1-ringer recruitment are 6.0 and 9.6 no $10^{9}$ for year classes 2004 and 2005 respectively, while the estimates for 0-ringers are 17.5, 27.8 and 11.9 no $10^{9}$ for year classes 2004, 2005 and 2006 respectively.

### 2.6 Assessment of North Sea herring

### 2.6.1 Data exploration and preliminary results

A benchmark assessment for North Sea herring was carried out in 2006. North Sea herring is on the AFCM observation list, but was also classed as an update assessment in 2007 by ACFM. With this in mind limited exploration was carried out into the fit of the assessment. The full choice of assessment model, catch and survey weightings and the length of separable period were not explored in detail in 2007.

### 2.6.1.1 Revision of historic data

Since last year's WG, a number of sources of data have been revised

1) The ICES IBTS database for the full time series (section 2.3.3??)

2 ) The acoustic survey data for 2005 (Section 2.3.1)
3 ) The catch age structure from The Netherlands in 2004 and 2005, catch in tonnes from UK in 2001-2004.

The 2006 assessment was rerun with the same settings with original and revised historic data. The differences were all less than $2 \%$, except for F in 2001-2004, which increased by an average of $5 \%$ and SSB in the terminal year (2005), which was reduced by $3.5 \%$.

### 2.6.1.2 Selection of indices and weighting in the assessment of North Sea herring

The usual assessment tool for the assessment of North Sea herring is ICA. The settings were the same as last year. Acoustic, Bottom trawl (IBTS), MIK and Larvae (MLAI) surveys are available for the assessment of North Sea autumn spawning herring. The surveys and the years for which they are available are given in Table 2.6.1.

The WG in 2003 made an extensive review covering both inverse variance and structural errors, and it considered that the inverse variance weighting method provided the better method. In 2006 the WG updated the variance weights and showed that the revised weights produced only a small change in the results. Following this examination it was decided that the weighting of surveys and catch is fixed between benchmark assessments as the sensitivity of the assessment to yearly revision of weights is small and the work required to do the analysis extensive. The weights express the WG view that the young herring are best estimated with MIK and IBTS surveys, the older herring are best evaluated through the acoustic survey and the SSB should be estimated through the MLAI.

The influence of individual surveys in terminal F and SSB is shown in Figure 2.6 .1 where the results of assessments based on catch data combined with each series one at a time are shown along with variance co-variance estimates of uncertainty in the terminal values. This shows that the MLAI alone would give a high value of SSB and low value of $F$, the other indices lie more within the cloud of points. Previous examination of the results of assessments using combinations of indices (ICES 2006) has shown that the best retrospective patterns are obtained when all the indices are used in the assessment.

### 2.6.1.3 Period of separable constraint

Changes in the regulations in 1996 have affected the various components of the fishery differently. During the period following these changes, meetings of this WG split the separable period into two different periods: 1992-1996 and 1997 onwards. In the WG 2001 it was considered that the number of years after the change in selection was long enough to use only a single separable period of four years. During 2002-2004 a separable period of five years was used. A retrospective study in 2002 found that year on year adaptation of the separable period did not improve the performance of the assessment model and that a fixed selection period gave more stable assessments, even with changing management. This year the WG noted that there was a small rise in selection at ages $5-7 \mathrm{wr}$. Therefore the WG investigated the effect of increasing selection at ages 8 and $9+$ to maintain a flat selection at old ages. No important differences in the model fit or outputs were detected. The estimation of F at reference age ( 4 wr ) was not significantly different and differences in estimates mean $\mathrm{F}_{2-6}$ and SSB in 2006 were found to be negligible. So the 5 year separable period with F at age 8 and $9+$ set equal to the reference age was maintained in the current assessment.

### 2.6.1.4 Model fit and residuals

The influence of the catch and the surveys was explored on the estimation of reference F and the model fit. ICA was run using all catch and survey data with the same procedure as last year. The patterns in catch residuals (Figure 2.6.2) are different to the assessments in 2006. The revision of data appears to have reduced the magnitude of the residuals overall by about $25 \%$ and particularly in the terminal year. There is no evidence of cohort effects across the full selection pattern but some residuals on cohorts are in a similar direction in adjacent years. Overall the catch residuals are small.

To explore the contribution of the catch and the survey data to the specific deviations in the data expressed in the residuals as large values, each data series was examined for large
residuals in the recent past. The individual data values were then removed and the influence of these points on the assessment evaluated. The following points were considered:
i ) Acoustic survey estimates of age 1 wr in 2005 and 2006,
ii ) IBTS estimates of 2 wr in 2006 and 2007,
iii ) MIK estimate of 0 wr in 2004
iv ) MLAI estimates of SSB in 2003 and 2004.
Table 2.6 .2 shows the influence of these points on the main management parameters mean F ages 2-6, SSB, TSB and recruitment. The age based surveys influence F, SSB and TSB by less than $3 \%$, the SSB index has a maximum influence of $6 \%$ on the same parameters. These changes are negligible both in the context of precision of the assessment and the management agreement. Changes to recruitment are similarly insignificant with the exception of the estimate of the 2004 year class. As this year class has only been included in catch figures at ages 0 and 1 wr , and these are down-weighted in the assessment, the surveys dominate the estimation of this parameter. There is considerable and unusually large conflict between IBTS and Acoustic survey estimates of the value of this parameter, with the MIK estimate lying between the others. The source data from both time series were examined independently of the assessment: the Acoustic survey gives a high estimate and has a higher than usual CV; in contrast the IBTS index gives a low value with much higher mortality age 1 to 2 wr than seen in any other pair of years. Thus both series appear potentially to have greater errors than would occur on average, so there is no obvious preference for one series over the other. The final estimate is somewhat in balance, the estimate being potentially pulled $+19 \%$ by one survey and $-33 \%$ by the other (Table 2.6.2). As it is impossible to assign preference, not further action is justified.

In previous assessments it has been noted that in recent years the MLAI has positive residuals, and the Acoustic survey has a block of negative residuals at older ages (Figure 2.6.2). The current assessment shows that this pattern has been maintained in history but the agreement in the terminal year appears to be better than that seen last year (Figure 2.6.2). In particular the 2006 residual in the MLAI is small. In the2006 benchmark assessment it was concluded that one of the reasons for the relatively stable assessment was the balance of three major sources of information, with each potentially delivering short periods with bias but in combination providing a balance of errors.

### 2.6.1.5 Analytical retrospective

Figure 2.6.3 shows retrospective estimates of mean F, SSB and recruitment, by removing one year of data at a time. The $F$ shows considerable consistency over the last 6 years, with underestimation during the period immediately following the management changes of 1996-7. SSB is more variable in the last 6 years showing upward and downward revision. The SSB has the same period of bias following the 1996-7 management changes as F. This retrospective analysis, which shows improvements over the analytical retrospective presented in 2006 (ICES 2006). This improvement suggests that the revision of catch data back to 2001 may have improved the data series. Through the use of FLR the retrospective analysis has been further extended this year to evaluate the retrospective influence of individual indices of SSB or surveys that include adults. These are used one at a time along with the catch data and MIK recruit index. This gives three retrospective analyses using MLAI, Acoustic and IBTS surveys as the main tuning fleets (Figures 2.6.4, 2.6.5 and 2.6.6 respectively). In all cases these show poorer retrospective patterns than the combined data set (Figure 2.6.3).

### 2.6.1.6 Conclusions of exploration of the assessment.

In 2006 the formulation of the assessment was supported by an extensive benchmark and it was judged a credible tool for management advice. The patterns in the residuals seen in previous years is still present though is less in catch and the terminal year of this assessment compared with the terminal year last year. This indicates that catch and survey indices show different signals, and this is confirmed when indices are fitted individually to the assessment (Figure 2.6.1). However, the overall balance of index information appears useful and the retrospective analysis suggests that currently the assessment using the full data set performs the best.

### 2.6.2 Final Assessment for NS herring

The final assessment of North Sea herring was carried out by fitting the integrated catch-atage model (ICA) with a separable constraint over a five-year period, tuned with the Acoustic survey (1989-2006), MLAI SSB index (1973-2006), IBTS (1984-2007) and the MIK survey (1992-2007) time series. The model settings are shown in Table 2.6.3, the ICA output is presented in Table 2.6.4, the stock summary in Table 2.6.5 and Figure 2.6.7 and model fit and parameter estimates in Table 2.6.8 and Figures 2.6.5-2.6.12.24

The spawning stock at spawning time in 2006 is estimated at approximately 1.2 million tonnes, declining from 1.6 million tonnes in 2005. The abundance of 0wr fish in 2006 (2005 year class) remains low for the fifth consecutive year and is currently estimated as the lowest since 1979. The strong 1998 and 2000 year classes are still evident in the population, with the 2000 year class at 5 wr in 2006 and the 1998 year class at 7 wr both being the highest in the time series since 1986/7. Mean fishing mortality on 2-6wr herring in 2006 is estimated at around 0.35 , which is above the management agreement $F$ of 0.25 , while mean $F$ on $0-1 \mathrm{wr}$ herring is 0.08 , below the agreed $F_{0-1}$ of 0.12 . The value of mean $F 2-6 \mathrm{wr}$ for 2005 in the current assessment is 0.37 , which is slightly higher than the value of mean $\mathrm{F} 2-6 \mathrm{wr}$ from last year's assessment, which was 0.35 . The SSB in 2005 has also been revised downwards by $6 \%$ from 1.7 to 1.6 million tonnes. The 2006 recruitment ( 0 group in 2007) is estimated at 12,000 million, which is $28 \%$ of geometric mean of recruitment since 1981.

### 2.7 Short term projection by fleets.

### 2.7.1 Method

The procedure and program used (MFSP Skagen; WD to HAWG 2003) was the same as has been used since 2003. For the North Sea herring, managers have agreed to constrain the total outtake at levels of fishing mortalities for ages $0-1$ and $2-6$, and need options to show the trade-off between fleets within those limits. The MFSP program was developed to cover these needs.

### 2.7.2 Input data

## Fleet Definitions

The current fleet definitions are:

## North Sea

Fleet A: Directed herring fisheries with purse seiners and trawlers. By-catches in industrial fisheries by Norway are included.

Fleet B: Herring taken as by-catch under EU regulations.

## Division IIIa

Fleet C: Directed herring fisheries with purse seiners and trawlers
Fleet D: By-catches of herring caught in the small-mesh fisheries
The fleet definitions are the same as last year.

## Input Data for Short Term Projections

All the input data for the short term projections are shown in Table 2.7.1, which is the input file for the predictions.

Stock Numbers: For the start of 2007 the stock numbers at age were taken from ICA (ica.n file)

Recruitment: For 2008 and 2009, the recruitment was set to 22963 million which is the geometric mean of the recruitments of the year classes 2001-2006, as estimated in this years assessment. This is less than half the mean recruitment used prior to 2006. The low recruitment was assumed because all the year classes from 2001 onwards have been poor. Analysis of the time series of SSB and recruitment data by the SGRECVAP (ICES CM 2006/LRC:03) clearly indicates a shift in the recruitment success in 2001. The underlying cause for the change in 2001 is not clear, but there is no evidence to justify an assumption of long term average recruitment from 2008 onwards. Consequently, the advice is adopted to the current recruitment regime.

Fishing Mortalities: Selection by fleet at age was calculated by splitting the total fishing mortality in 2006 at each age proportional to the catches by fleets at that age (Table 2.2.6). These fishing mortalities were used for all years in the prediction.

Mean weights in the catch by fleet: The mean weights by fleet for the years 2004 - 2006, excluding the 2000 year class, were used for all year. For the 2000 year class, the weights at age in the catches by the A-fleet have so far been in the order of $10 \%$ below the average of the adjacent year classes, and the difference appears to be increasing. Assuming that the 2000 year class will continue to have reduced weights at age, the weights at age for this year class were reduced in the prediction years. This was done by fitting a second order polynomial to the weights at age observed so far and extrapolating this function to the prediction years. The resulting weights for the 2000 year class in 2007 and 2008 are about $12 \%$ lower than the corresponding weights at age used for other year classes. For the fleets B, C and D, no adjustments were made. The lower weight at age of the 2000 year class has not been apparent in the catches of these fleets. For the C and D fleets, the reason may be that the samples for weight at age are likely to contain both autumn spawners and spring spawners. For the B-fleet, this year class is hardly represented in the catches any more.

Mean Weights at age in the stock: The smoothed weights at age in the stock for 2006 were used. However, the weights at age for the 2000 year class were reduced by $12 \%$, which is the same reduction as for weights in the catch.

Maturity at age: The average maturity at age for 2004 to 2006, calculated without the 2000 year class, was used (Table 2.6.2.2). The 2000 year class is now fully mature.

Natural Mortality: Unchanged from last year, equal to those assumed in the assessment.
Proportion of $\mathbf{M}$ and $\mathbf{F}$ before spawning: Unchanged from last year at 0.67 .

### 2.7.3 Prediction for 2006 and management option tables for 2007

### 2.7.3.1 Assumptions for 2007

After the TACs were increased in 2003, the TAC for the A-fleet has been over-fished by 9 16 percent, while the other fleets caught less than half their TAC or by-catch ceiling. Catches in 2007 may be predicted with some confidence. The retrospective error has been low in recent years. It therefore seems most reasonable to use assumed catches to account for the removal in 2007.

In previous years it has been assumed that the TAC for the A-fleet would be overshot as before. The overshoot has gone down in the most recent years ( $17 \%$ in 2004, $15 \%$ in 2005, $10 \%$ in 2006). Therefore, it is assumed that the TAC of 341000 tonnes for 2007 for the Afleet will be overshoot by $10 \%$, which is the overshoot in 2006.

The utilisation of the by-catch quota by the B-fleet has fluctuated between $23 \%$ and $44 \%$ since 2003, and was $28 \%$ in 2006. For the prediction, it is assumed that $33 \%$ of the bycatch-quota will be taken, which is the average percentage since 2003. For the $C$ and $D$ fleet, it was assumed that their catch of North Sea autumn spawning herring would be the same as in 2006. The fishing mortalities resulting from these assumed catches were close to the fishing mortalities by fleet for 2006. Thus the difference between a catch constraint and F status quo constraint for 2007 therefore is small.

### 2.7.3.2 Management Option Tables for 2008

The EU-Norway agreement on management of North Sea herring was updated in 2004. The revised rule specifies fishing mortalities for juveniles ( $\mathrm{F}_{0-1}$ ) and for adults ( $\mathrm{F}_{2-6}$ ) not to be exceeded, at 0.12 and 0.25 respectively, for the situation where the SSB is above 1.3 million tonnes. In addition, it now has a rule specifying reduced fishing mortalities when the SSB is below 1.3 million tonnes. Moreover, the current agreement has a constraint on year-to-year change of $15 \%$ in TAC, but allows for a stronger reduction in TAC if necessary.

The rule for reducing $F$ at $\mathrm{SSB}<1300$ thousand tonnes derives the F from the SSB as
$\mathrm{F}_{2-6}=0.25-\left(0.15^{*}(1300-\mathrm{SSB}) / 500\right)$
$\mathrm{F}_{0-1}=0.12-(0.08 *(1300-\mathrm{SSB}) / 500)$
The interpretation by HAWG is that the SSB referred to should be the SSB in the prediction year, i.e. the Fs for 2008 should reflect its consequence for SSB in 2008.

Because of the recent poor recruitment, the management rule above leads to a strong reduction in quotas. The management agreement has a $15 \%$ limit on the change in TAC from one year to the next, and a clause to abandon this limit if needed. Short term predictions are presented for both alternatives.

With four fleets there are innumerable combinations of fleet-wise fishing mortalities and catches that satisfy the agreed rules.

Since the North Sea autumn spawning (NSAS) stock was rebuilt, the advice has been that the primary limiting factor for the fishery in IIIa should be the concern for the Western Baltic spring spawning (WBSS) stock. Due to the low recruitment og NSAS in particular, but also of WBSS, more restrictive quotas for the IIIa may be necessary. Using that as a guideline, options for catches by the fleets C and D were derived from two options for the outtake of WBSS:

1) Assuming a total WBSS catch of 76.3 thousand tonnes, representing a $15 \%$ reduction in TAC, gives NSAS catches for the C and D fleets of 14.4 and 7.6 thousand tonnes respectively
2 ) Assuming a total WBSS catch of 40.2 thousand tonnes, corresponding to a fishing mortality for WBSS at $\mathrm{F}_{0.1}=0.22$, gives NSAS catches for the C and D fleets of 6.9 and 3.7 thousand tonnes respectively.

For 2007, Norway is allowed to transfer $40 \%$ of its quota in IIIa to IV. To show the effect on the stock of this transfer, which amounts to 3820 tonnes in 2007, an example was made where this catch was added to the A-fleet for 2007. The expected fraction of this catch that would be NSAS (720 tonnes) was subtracted from the C-fleet for 2007.

The following options for 2008 are tabulated:

1) Following the harvest rule without constraints on the year-to-year change in catch, assuming catches by the C and D fleets corresponding to a $15 \%$ reduction in TAC for WBSS.
2 ) Following the harvest rule without constraints on the year-to-year change in catch, assuming catches by the C and D fleets corresponding to $\mathrm{F}_{0.1}$ for WBSS
3 ) As 1 , but with a catch by the A-fleet of 289800 tonnes, which is $85 \%$ of the TAC for 2007.

4 ) As 2, but with a catch by the A-fleet of 289800 tonnes, which is $85 \%$ of the TAC for 2007.

5 ) As 1, but with a transfer of 3820 tonnes of catch from the C-fleet to the A-fleet.
6 ) No fishing
All predictions are for North Sea autumn spawning herring only.
The results are presented in Table 2.7.2.

### 2.7.4 Comments on the short-term projections

The outlook for this stock is poorer than in previous years, due to the recent reduction in the recruitment. This has been taken into account in the current prediction, both through the stock numbers at the start of 2007 as derived from the assessment, and by assuming a recruitment in line with what has been experienced the last 5 years.

As a result, even without fishing, the SSB will be well below 1.3 million tonnes in 2008. The present agreement includes a rule to reduce the fishing mortality below 0.25 if the SSB is below 1.3 million tonnes, but with the option to limit the reduction in TAC to15\%. In the present situation, applying the $15 \%$ rule will lead to an $\mathrm{F}_{2-6}$ well above the agreed 0.25 . Moreover, medium term predictions indicate a substantial risk to $\mathrm{B}_{\mathrm{lim}}$ if the catches are reduced this slowly.

The effect of the assumed reduction in catches in IIIa is small in the short term, but will be more important in the medium term as the fishery in IIIa mostly exploits juveniles. Likewise, it is assumed that fishing mortality for the B-fleet follows the management rule, implying a continued low exploitation of juveniles in the North Sea,

The predictions presented here account for the slow growth of the large 2000 year class. There are no indications of reduced growth of the subsequent year classes.

The estimated impact of the juvenile fishery depends on the assumed value for natural mortality. It has not been investigated to what extent changes in natural mortality would affect the current advise, or if indeed such changes are taking place. However, some of the important predator stocks are currently in a poor condition.

### 2.8 Medium term predictions and HCR simulations

Medium term predictions have been made repeatedly for many years for this stock, to first develop and later evaluate the current management agreement. In all these simulations, a recruitment in line with what has been experienced in the 40 years in the past was assumed. This assumption may now be questioned, given the poor recruitment for 6 consecutive years. Such a prolonged sequence of poor recruitments is unusual for this stock, and would appear as very unlikely in the previous simulations. SGRECVAP has considered variability in the timeseries of recruitment and concluded that at least from a statistical point of view the current situation is different from the past. Although the cause of this poor recruitment is unclear and it is uncertain whether it will continue, management may have to adapt to a lower productivity of the North Sea herring stock.

To inform management under these circumstances medium term predictions assuming low recruitment in the future were presented in last years report. These studies are further extended here. This is done to give some guidance to management adaptation to a reduced productivity

The software used was STPR3, the same as used at the evaluation of HCRs for North Sea herring in June 2004. This is a program for performing 10 years stochastic simulations of the stock and fishery, applying some HCRs. A description can be found i.a. the SGMAS report (ICES CM 2005 /ACFM:09) and a manual in an EU Norway report on medium term management measures (EU 2004).

### 2.8.1 Input data

The program was run with 2 fleets, Fleet 1 corresponds to the A-fleet and Fleet 2 corresponds to fleets $\mathrm{B}, \mathrm{C}$ and D combined.

Stock numbers in the initial year 2007 and their variances-covariances were taken from the current ICA output (ica.n and ica.vc). The stock-recruitment function was the same as used in previous simulations ('Ockhams razor’), but with a reduced recruitment. It assumed recruitment of 22963 millions independent of SSB at SSB larger than 800 thousand tonnes, and a linear reduction of the recruitment at lower SSB. The recruitment was drawn from a lognormal distribution with $\sigma=0.35$. These values are the mean and standard deviation (on a log scale) in the recent 6 years recruitment data series.

For weights and maturities historical data were used, by drawing years randomly and using data from that year.

Fleetwise selection at age were equal to those used in the short term prediction (Table 2.7.1)
For the intermediate year, catches by fleets were assumed as in the short term prediction.
Assessment was assumed to deviate from the true values by a random multiplier with mean 1.1 and $\mathrm{CV}=0.1$. For implementation error, a CV of 0.1 was assumed throughout.

To show the sensitivity to some of these assumptions, one run (run 4) was made with a CV on recruitment of 0.58 (as in the full historic series), and one (run 5) with a breakpoint in the stock-recruitment function at 500. Also, the risk to Blim in the last year (2017) for a range of levels of implementation error is shown, all assuming a CV on that error of 0.1.

### 2.8.2 Simulation options.

- Run 1 Standard HCR: The first set of simulations applied the basic harvest rule agreed by Norway and EU from 2004:

At $\mathrm{SSB}>1.3$ million tonnes: $F 0-1=0.12$ and $F 2-6=0.25$
At $S S B<1.3$ million tonnes and SSB $>800000$ tonnes:

$$
\begin{aligned}
& F 0-1=0.12-\left(0.08^{*}(1300000-S S B) / 500000\right) \\
& F 2-6=0.25-(0.15 *(1300000-S S B) / 500000) 800000 \text { tonnes: }
\end{aligned}
$$

For SSB $<800000$ tonnes: $F 0-1=0.04$ and F2-6 $=0.10$
The agreement does not state the year which the SSB refers to. The SSB considered by STPR3 is the SSB in the quota year.

- Run $2-15 \%$ rule: The second set applied the rule to not change the TAC by more than $15 \%$ per year. The other parameters were as in the first set.
- Run 3 - Applying fishing mortalities F0-1 = 0.12 and F2-6 $=0.25$ : The other parameters were as in run 1, i.e. the rule constraining catch variation was not applied.
- Run 4 - as run 2 (i.e. with the $15 \%$ rule), but with a larger CV $=0.58$ on recruitment
- Run 5 - as run 2 (i.e. with the $15 \%$ rule), but with a lower breakpoint (= 500000 tonnes) in the stock-recruit relation.


### 2.8.3 Results

The main results for each run are shown in Figures 2.8.1-5. The risk associated with implementation error is shown in Figure 2.8.6

Run 1 shows that with the harvest rule implemented with no error, the risk to Blim is small. The SSB settles slightly above 1 million tonnes, and the catches in the order of 2-3 hundred thousand tonnes. With the $15 \%$ rule (Run 2), there is a considerable risk to Blim around 2010, with a fair chance of recovery once the fishing mortality has come down towards 0.2 . Compared to Run 1, the catches will be higher in the first years, but lower thereafter. With the Fs of 0.12 and 0.25 , the risk to Blim is substantial and stable. The assumptions about recruitment variation and breakpoint in the stock-recruit function do not appear to have a major impact on the results.

As shown in Figure 2.8.6, the current management rule is not robust to implementation error in terms of overfishing of the quotas beyond the level that is estimated at present. Hence, either the enforcement has to be stronger than in the past, or an even more conservative harvest rule has to be applied in order to safeguard against depleting the stock.

The present simulations have been done with weights and maturities representing the whole historic time series. The weights in recent years have tended to be lower that previously, which may explain why the predictions for 2007 give a somewhat larger biomass than the short term prediction. There is, however, no firm basis for assuming lower growth in the future.

### 2.9 Precautionary and Limit Reference Points

In 2003, SGPRP (ICES 2003 ACFM:04) suggested to reduce $B_{\text {lim }}$ from the current 800000 tonnes to about 560000 tonnes, based on the results of the segmented regression analysis of the stock and recruitment data. Fitting a segmented regression stock-recruit function with nonlinear minimisation of the SSQ of log residuals suggests a break point at 537000 tonnes.

In 2007 WKREF explored limit reference points for NSAS herring among a number of other stocks. WKREF concluded that there is no basis for changing $\mathrm{B}_{\mathrm{lim}}$ based on this analysis. SGRECVAP results could be basis for revisiting reference points. The distance between a management reference point (trigger or $\mathrm{B}_{\mathrm{pa}}$ ) and $\mathrm{B}_{\text {lim }}$ defines a risk and should be evaluated in the context of harvest control rules in consultation with stakeholders and managers.

HAWG decided not to propose any revision of the $\mathrm{B}_{\mathrm{lim}}$ reference points at present for the following reasons:

- WKREF questioned the validity of the current calculation procedure for the segmented regression.
- Currently there is indications that the stock dynamics are changing
- The role of regime shifts in determining limit reference points should be integrated in the process
- HAWG would prefer to consider all reference points together, rather than revising just $\mathrm{B}_{\text {lim }}$.

Most importantly, a downward revision of reference points now would not be helpful in precautionary management of the stock. When properly applied the harvest control rule in place for this stock has worked well in the recent past, and apart from $\mathrm{B}_{\mathrm{lim}}$, the current reference points are derived from this HCR. The target F in the HCR was adopted by ACFM as $\mathrm{F}_{\mathrm{pa}}$, while the trigger point at which F should be reduced below the target is adopted as $\mathrm{B}_{\mathrm{pa}}$.

### 2.10 Quality of the Assessment

### 2.10.1 Precision of historic timeseries

A bootstrap variance covariance evaluation of the precision of the assessment carried out using ICA is shown in Figure 2.10.1. The historic uncertainty supports the view that the stock has declined in recent years, and that recruitment is lower than since the late 1970s.

### 2.10.2 Comparison with earlier assessments

The 2007 assessment is in good agreement with last years assessment and the intermediate year in the short term projection, see table below.

| Assessment year | SSB in 2005 | F2-6 in 2005 | SSB in 2006 | F IN 2006 |
| :--- | :--- | :--- | :--- | :--- |
| 2007 | 1.59 M t | 0.37 | Assessed 1.21 Mt | Assessed 0.35 |
| 2006 | 1.69 M t | 0.35 | Projected 1.33 Mt | Projected 0.35 |

There has been a downward revision of SSB by about $6 \%$ of which about half is due to revision of catch (Section 2.1 and 2.6) but F is estimated to be very similar.

Cohort retrospectives are shown in Figure 2.10.2. The earliest cohorts shown have some revision over the early years. Latterly the cohort retrospective evaluations suggest the WG is providing a fairly consistent evaluation of most year classes. The exceptions are 2001 and 2004 year classes. In particular the dominant 2000 year class has been estimated consistently since it was first seen in 2001.

The both assessment and projections currently appear to be a good basis for management advice.

### 2.11 Herring in Division IVc and VIId (Downs Herring).

Over many years the working group has attempted to assess the contribution of winter spawning Downs herring to the overall population of North Sea herring. Since 1985, there is a separate TAC for herring in Divisions IVc and VIId as part of the total North Sea TAC.

Historically, the TAC for herring in IVc and VIId has been set as a proportion of the total North Sea TAC and this has varied between 6 and $16 \%$ since 1986. The proportion has been relatively high in recent years, particularly since 2002. However, ACFM in 2005 expressed a range of concerns regarding Downs herring and recommended that the proportion used to determine the TAC should be set to the long term average of the proportions used since 1986 (11\%). In accordance with ICES advice the sub-TAC was reduced from 74000 tonnes in 2005 to 50023 tonnes in 2006 (a reduction of $33 \%$ compared with 2005). For 2007, the same proportion (11\%) was kept and the TAC was set at 37517 tonnes. (Figure 2.11.1).

ACFM has in the past expressed concern that there is a persistent tendency to overfish the Downs TAC. However, this tendency has been markedly reduced in recent years (Figure 2.11.2), possibly because the TACs have been much higher. Landings in 2006 amounted to 56 597 tonnes, slightly higher than the TAC.

A further concern is that recent high catch levels in IVc and VIId have been driven largely by the strong 2000 year class. This year class accounted for $67 \%$ and $51 \%$ of the catch in numbers in 2004 and 2005 respectively but has reduced to $37 \%$ in 2006. As has been noted previously these fish are smaller and less mature than the average for a given age therefore, if the fishery preferentially takes lighter fish the resulting F is comparatively higher.

Historically, the Downs herring has been considered highly sensitive to overexploitation (Burd, 1985; Cushing 1968; 1992). It is less fecund and expresses different growth dynamics and recruitment patterns to the more northern spawning components. Furthermore, the directed fishery in Q4 and Q1 targets aggregations of spawning herring. Preliminary studies undertaken by this WG in 2006 (ICES CM 2006) based on population profiles suggested that total mortality (Z) was significantly higher for the 1998 and 1999 year classes of Downs herring compared to other classes caught in the Northern part.

Downs herring is also taken in other herring fisheries in the North Sea. Downs herring mixes with other components of North Sea herring in the summer whilst feeding, but it has not been possible to quantify the Downs component in the catch. There is also a summer industrial fishery in the eastern North Sea exploiting Downs and North Sea autumn spawning herring juveniles. Tagging experiments in the Eastern North sea (Aasen et al, 1962) estimated that around $15 \%$ of those catches comprised Downs recruits. Otolith microstructure studies of catches from the northern North Sea suggested that the proportion of Downs herring may vary considerably from year to year ( 26 to $60 \%$ ) and may also vary between fleets (Dickey-Collas et al., 2005).

The proportion of the autumn and winter spawning components in recruiting year classes of North Sea herring has been traditionally monitored through the abundance of different sized fish in the IBTS. The 1-ring fish from Downs spawning sites (winter) are thought to be smaller than those from the more northern, autumn spawning sites ( $<13 \mathrm{~cm}$ and $>13 \mathrm{~cm}$ respectively). Both the total abundance and the proportion of Downs herring have, on average, been comparatively higher since the early 1990s, although there is considerable variation between year classes (Figure 2.11.3, Table 2.3.3.3). These data suggest that around $35 \%$ of the strong 2000 year class came from Downs production and that approximately $70 \%$ of the 2002 year class originated from Downs production. The percentage contribution of the 2005 year class is about the same as the long-term average $=23 \%$ and appears to be stronger than the 2004 year class (Fig. 2.11.3).

2006 year class - The recruitment for the 2006 year class (Figure 2.11.4, MIK index) appears lower than the 2005 year class. With the extension of the IBTS area in the Eastern English Channel, the number of MIK samples, distributed in 4 statistical rectangles, increased during the last survey and therefore, results are considered more reliable.

Last year the EC set a proportion of TAC for herring in IVc and VIId in accordance with the ICES advice. The TAC is specific to the conservation of the spawning aggregation of Downs herring. In the absence of other information there are uncertainties in the recruitment to the component in the next few years and HAWG recommends that the IVc-VIId TAC should be maintained in 2008 at $11 \%$ of the total North Sea TAC (as recommended by ACFM). This recommendation should be seen as an interim measure prior to the development of a more robust harvest control rule for setting the TAC of Downs herring, supported by increased research effort into the dynamics of this component in fisheries in the central and northern North Sea. Any new approach should provide an appropriate balance of F across stock
components and be similarly conservative until the uncertainty in the Downs contribution to the catch in all fisheries in the North Sea is reduced.

## Extension of the IBTS area and acoustic survey in the Eastern Channel.

Winter spawning Downs herring stock is exploited off the eastern English Channel by different fleets, mainly at the end of the year. The rest of the year, this stock component is mixed with the overall population of North Sea herring in the feeding grounds. According to French fishermen this pattern seems to have changed as large herring shoals have been observed in the English Channel, mainly along French coasts, until April.

In March 2006, after the IBTS survey, the French RV "Thalassa" recorded acoustics data that confirmed the fishermen's observations. Shoals of significant size were observed in coastal waters in the ICES rectangle 30F1. Some trawl hauls were made and the catches consisted of herring with mean length of 25 cm .

During the last IBTS WG in March 2006, the extension of the IBTS $1^{\text {st }}$ quarter survey area in the Eastern English Channel was considered: additional GOV hauls carried out in this area would provide more information on Downs herring and its distribution at this period of the year. The IBTS WG agreed that RV "Thalassa" could take some additional trawl hauls when it started its IBTS cruise at the end of January on its way through the English Channel before going to the North sea. The HAWG supported the idea and the extension of the IBTS area was implemented at the 1st quarter IBTS 2007.

During 4 days, (30 January - 2 February 2007) the RV "Thalassa" covered the Eastern part of the English Channel. 8 GOV hauls and 20 MIK stations were made in each ICES rectangle according to the IBTS protocol. In addition, acoustic data were recorded during day and night and 5 pelagic hauls made when fish marks were detected. Because of the vessel traffic in this area, the lack of time to do a full coverage and the impossibility to cross sandbanks, only two acoustic transects were done along the English and the French coasts and a third one in the middle of the English Channel (Figure 2.11.5).

The most important marks were recorded along French coasts and the catch composition of pelagics hauls consisted of mixed herring, sardine and other pelagic species in the south of the area and mainly herring in the northern part. For herring; the catch composition consisted of 26 cm mean length fish belonging to age-groups 3 - 6. (Figure 2.11.6).

According to fishermen's observations, very large and continuous shoals of herring were found at the same time as this survey in a local area, concentrated along sandbanks and observed again when IBTS was finishing at the end of February. Mean density could be estimated of between 500 and 1500 tonnes per nautical mile square but it could not be raised to the whole area due to the spatial heterogeneity and the sampling protocol used.

As it was the first year that the survey was carried out in this area at this time of the year, it must be considered preliminary. Further, the survey design needs to be improved and pelagic samples need to be increased. Though a reliable biomass estimation on herring during its migration through the English Channel is likely to be difficult, the survey could certainly give more information on herring shoals observed, their evolution and the possible change in behaviour in relation to herring spawning area.

In the 2006 HAWG report, some rules to set the percentage southern North Sea TAC allocation were proposed but no simulation testing of those was performed. The following pieces of information are required to develop the analysis further:

- catch at age by area,
- microincrement analysis of otoliths (to determine spawning type), expanded to other fleets in the North Sea, high resolution MIK coverage in Southern North Sea and the Channel area
- extension of IBTS in the Eastern Channel.

Hence, HAWG continues to recommend that existing surveys of herring in the southern North Sea and English Channel be maintained and that the microincrement analysis of otoliths currently undertaken are continued and expanded to other fleets in the North Sea. Further, extensive simulation testing of alternative HCRs needs to take place.

### 2.12 Management Considerations

Based on the most recent estimates of SSB and fishing mortality, the North Sea autumn spawning herring stock is considered to be at 1.2 million t in 2006 and is expected to decrease to 0.97 million tonnes in 2007. F in 2006 was 0.35 and expected to be similar at $\mathrm{F}=0.33$ in 2007. Following currently estimated low recruitment, SSB is expected to remain at about this level of biomass for a while, declining further or rising slowly depending on the level of F .

SSB peaked after the rise from the low stock size in the mid-1990s, in response to reduced catches, strong recruitment and management measures that reduced exploitation both on juveniles and adults. However, in the last 5 years the recruitment has been at $40 \%$ average, and the stock is declining. Landings of adult herring in recent years have consistently exceeded the agreed TAC, mainly due to unallocated catches and catches misreported out of the North Sea (see section 2.1). The fishing mortality has increased, mainly due to the management rule that limits reduction to $15 \%$ per year, and at 0.35 is now above what was intended in the management agreement, and what was considered sustainable. If F is maintained at this level SSB will decline slowly over the next few years and may reach Blim in 2009 or 2010.

The stock is managed according to the EU-Norway Management agreement which was updated on 26 November 2004, the relevant parts of the text are included here for reference:

1. Every effort shall be made to maintain a level of Spawning Stock Biomass (SSB) greater than the 800,000 tonnes (Blim).
2. Where the SSB is estimated to be above 1.3 million tonnes the Parties agree to set quotas for the directed fishery and for by-catches in other fisheries, reflecting a fishing mortality rate of no more than 0.25 for 2 ringers and older and no more than 0.12 for 0-1 ringers.
3. Where the SSB is estimated to be below 1.3 million tonnes but above 800,000 tonnes, the Parties agree to set quotas for the direct fishery and for by-catches in other fisheries, reflecting a fishing mortality rate equal to:

$$
\begin{aligned}
& 0.25-\left(0.15^{*}(1,300,000-S S B) / 500,000\right) \text { for } 2 \text { ringers and older, and } \\
& 0.12-\left(0.08^{*}(1,300,000-S S B) / 500,000\right) \text { for } 0-1 \text { ringers. }
\end{aligned}
$$

4. Where the SSB is estimated to be below 800,000 tonnes the Parties agree to set quotas for the directed fishery and for by-catches in other fisheries, reflecting a fishing mortality rate of less than 0.1 for 2 ringers and older and less than 0.04 for 0-1ringers.
5. Where the rules in paragraphs 2 and 3 would lead to a TAC which deviates by more than 15\% from the TAC of the preceding year the Parties shall fix a TAC that is no more than $15 \%$ greater or $15 \%$ less than the TAC of the preceding year.
6. Not withstanding paragraph 5 the Parties may, where considered appropriate, reduce the TAC by more than $15 \%$ compared to theTAC of the preceding year.
7. By-catches of herring may only be landed in ports where adequate sampling schemes to effectively monitor the landings have been set up. All catches landed shall be deducted from the respective quotas set, and the fisheries shall be stopped immediately in the event that the quotas are exhausted
8. The allocation of TAC for the directed fishery for herring shall be $29 \%$ to Norway and $71 \%$ to the Community. The by-catch quota for herring shall be allocated to the Community
9. A review of this arrangement shall take place no later than 31 December 2007.
10. This arrangement enters in to force on 1 January 2005.

ACFM examined the performance of this revised harvest control rule in 2005, and considered "the agreement in terms of target F to be consistent with the Precautionary Approach. However, ACFM also considered that the strict application of the TAC change limit of $15 \%$ (rule number 5) is not consistent with the Precautionary Approach in a situation like the present when five consecutive weak year classes have recruited to the population. The harvest control rule is in accordance with the precautionary approach if paragraph 6 is consistently invoked sufficiently early to prevent or minimise the risk of SSB falling below $\mathrm{B}_{\mathrm{pa}}$ even in the case of several consecutive weak year classes. Assuming that paragraph 6 would be invoked when TAC constraints would lead to SSB falling below $B_{p a}$ it is considered that the revised HCR is in accordance with the Precautionary Approach."

The situation now is unusual, and had not been anticipated, with all the five year classes from 2002 onwards being poor. The SGRECVAP, which was set up to have a closer look at the recruitment failure in herring (as well as in Sandeel and Norway pout), concluded that the reduced recruitment is caused by an increased mortality in the first winter. Analysis of the time series of SSB and recruitment data clearly indicates a shift in the recruitment success in 2001. An analysis of stock production (Section 1.8) shows similar results. The underlying cause for the change in 2001 is not clear, but there is no evidence to justify an assumption of normal recruitment from 2008 onwards.

Given the current sustained low level of recruitment, considering the $\mathrm{B}_{\text {trig }}$ in the management plan as $\mathrm{B}_{\mathrm{pa}}$ may be unrealistic and it is preferable to evaluate the precautionary nature of the management plan as a whole rather than referring to a biomass reference point that may not be achievable.

Following evaluation, the agreed plan is considered precautionary and the risk of SSB falling below $\mathrm{B}_{\mathrm{lim}}$ in the medium term is less than $5 \%$, when:

- current low levels of recruitment continue,
- implementation is constrained to give less than $10 \%$ over exploitation,
- there are no year on year restrictions on change in TAC

The continuation of the $15 \%$ year on year restriction in change in TAC increases the risk of SSB falling below $\mathrm{B}_{\mathrm{lim}}$ to greater than $25 \%$ over the next 5 years. Alternatively an implementation error of $30 \%$ will have a similar effect. Implementation error is currently estimated at just under $10 \%$ for the human consumption fleet in 2006, though it has been higher in the past.

The failure to comply with precautionary management rule in setting the TAC in 2007 has given rise to a fishing mortality that is higher than was envisaged. This is a matter of concern in a situation of extreme low recruitment. The consequences of the maintaining the present
fishing mortality at around 0.35 , has not been examined in detail, but it is clear that if it is not reduced it will lead to a substantial reduction in SSB to a level below $\mathrm{B}_{\mathrm{lim}}$ in the near future.

Consequently, the WG considers that the advice for 2008 should be adapted to the current recruitment. by allowing necessary year on year change in TAC and complying with the F reduction in the management plan and ensuring implementation of regulation is to better than $10 \%$.

This stock complex also includes Downs herring (herring in Divisions IVc and VIId), which has shown independent trends in exploitation rate and recruitment, but cannot be assessed separately. This year the Working Group concludes that the current state of the component is unknown. The WG's understanding of the component's dynamics is unlikely to improve until further examination of catch and the existing time series of surveys takes place. Both, alternative assessment methods have to be explored, and a greater knowledge of the ecology of Downs herring is needed. The Downs fishery is concentrated on the spawning aggregations in a restricted area, which makes this stock component particularly vulnerable to excessive fishing pressure. Catches of the Downs component are taken both in the southern area and in the mixed fishery in the central and northern North Sea. The EU splits its share of the total North Sea herring TAC (Subarea IV and Division VIId) into TACs for Divisions IVa+IVb and for Divisions IVc+VIId. ICES has proposed that a share of $11 \%$ on the total North Sea TAC (average share 1989-2002) would be appropriate for distributing the harvesting among Downs Herring and other stock components. While the WG acknowledges that the basis for this exact $11 \%$ figure is weak there are strong indications that the total mortality on the Downs component, of which fishing is the major component, has recently been significantly higher than for the rest of the NS components.

For the last few years since the North Sea autumn spawning (NSAS) stock was rebuilt, the ICES advice has been that the primary limiting factor for the fishery in IIIa should be the concern for the Western Baltic spring spawning (WBSS) stock. With an expected decline of the NS herring below 1.0 million t in 2007 primacy of consideration must be given to protection of this stock. The provision of advice for the NS affects the C and D fleets operating in IIIa. Projections for the WBSS stock also indicate poor recruitment and an expected decline in SSB with present F levels, and an incremental reduction of fishing mortality towards F0.1 is therefore advisable for this stock. This issue is dealt with in detail in the discussion of short term predictions in Section 2.7. and in Section 3.10. In should be noted that in setting the catch of WBSS herring the corresponding catch of NSAS herring in the D fleet puts specific restraints on the catches of the B-fleet, and some options may be rather restrictive.

All of the relaxation of area and bycatch rules are now contributing to the increase in exploitation and current over exploitation of North Sea herring. Removal of these derogations and increased compliance would be beneficial, particularly in the current circumstance of a declining North Sea population.

Table 2.1.1: Herring caught in the North Sea (Sub-area IV and Division VIId). Catch in tonnes by country, 1997 - 2006. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.

| Country | 1997 | 9 | 1998 | 9 | 1999 | 9 | 2000 |  | 2001 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 1 |  | - |  | 2 |  | - |  | - |  |
| Denmark | 38324 |  | 58924 |  | 61268 |  | 64123 |  | 67096 |  |
| Faroe Islands | 1156 |  | 1246 |  | 1977 |  | 915 |  | 1082 |  |
| France | 14525 |  | 20784 |  | 26962 |  | 20952 |  | 24880 | 14 |
| Germany, Fed.Rep | 13380 |  | 22259 |  | 26764 |  | 26687 |  | 29779 |  |
| Netherlands | 35985 |  | 49933 |  | 54467 |  | 54341 |  | 51293 |  |
| Norway ${ }^{4}$ | 41606 |  | 70981 |  | 74071 |  | 72072 |  | 75886 |  |
| Poland |  |  |  |  |  |  |  |  | - |  |
| Sweden | 2253 |  | 3221 |  | 3241 |  | 3046 |  | 3695 |  |
| USSR/Russia | 1619 |  | 452 |  | - |  | - |  | - |  |
| UK (England) | 3470 |  | 7635 |  | 11434 |  | 11179 |  | 14582 |  |
| UK (Scotland) | 22582 |  | 31313 |  | 29911 |  | 30033 |  | 26719 |  |
| UK (N.Ireland) | - |  | 1015 |  | - |  | 996 |  | 1018 |  |
| Unallocated landings | 63403 | 6,12 | 70329 | ${ }^{12}$ | 43327 | ${ }^{12}$ | 61673 | 12 | 27362 | ${ }^{12}$ |
| Total landings | 238304 |  | 338092 |  | 333424 |  | 346017 |  | 323392 | ${ }^{14}$ |
| Discards | - |  | - |  | - |  | - |  | - |  |
| Total catch | 238304 |  | 338092 |  | 333424 |  | 346017 |  | 323392 | 4 |
| Estimates of the parts of the catches which have been allocated to spring spawning stocks |  |  |  |  |  |  |  |  |  |  |
| IIIa type (WBSS) | 979 |  | 7833 |  | 4732 |  | 6649 |  | 6449 |  |
| Thames estuary ${ }^{5}$ | 202 |  | 88 |  | 88 |  | 76 |  | 107 |  |
| Others ${ }^{11}$ | - |  | - |  | - |  | 378 |  | 1097 |  |
| Norw. Spring Spawners ${ }^{\text {T }}$ | 54728 |  | 29220 |  | 32106 |  | 25678 |  | 7108 |  |


| Country | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 23 | 5 | 8 | 6 | 3 |
| Denmark ${ }^{7}$ | 70825 | 78606 | 99037 | 128380 | 102322 |
| Faroe Islands | 1413 | 627 | 402 | 738 | 1785 |
| France | 25422 | 31544 | 34521 | 38829 | 49475 |
| Germany | 27213 | 43953 | 41858 | 46555 | 40414 |
| Netherlands | 55257 | 81108 | 96162 | 81531 | 76315 |
| Norway ${ }^{4}$ | 74974 | 112481 | 137638 | 156802 | 135361 |
| Poland | - | - | - | 458 |  |
| Sweden | 3418 | 4781 | 5692 | 13464 | 10529 |
| Russia | - | - | - | 99 | - |
| UK (England) | 13757 | 18639 | 20855 | 25311 | 22198 |
| UK (Scotland) | 30926 | 40292 | 45331 | 73227 | 48428 |
| UK (N.Ireland) | 944 | 2010 | 2656 | 2912 | 3531 |
| Unallocated landings | 31552 | 31875 | 48898 | 57788 | 18764 |
| Total landings | 335724 | 445921 | 533058 | 626101 | 509125 |
| Discards | 17093 | 4125 | 17059 | 12824 | 1492 |
| Total catch | 352817 | 450046 | 550117 | 638925 | 510617 |
| Estimates of the parts of the catches which have been allocated to spring spawning stocks |  |  |  |  |  |
| IIIa type (WBSS) | 6652 | 2821 | 7079 | 7039 | 10954 |
| Thames estuary ${ }^{5}$ | 60 | 84 | 62 | 74 | 65 |
| Others ${ }^{11}$ | 0 | 308 | 0 | 0 | 0 |
| Norw. Spring Spawners ${ }^{15}$ | 4069 | 979 | 452 | 417 | 626 |

${ }^{4}$ Catches of Norwegian spring spawners removed (taken under a separate TAC)
${ }^{5}$ Landings from the Thames estuary area are included in the North Sea catch figure for the UK (England)
${ }^{7}$ Including any by-catches in the industrial fishery
${ }^{9}$ Figures verified and altered if needed in 2003 by SG Rednose (ICES 2003/ACFM:10)
${ }^{10}$ Figure altered in 2001
${ }^{11}$ Caught in the whole North Sea, partly included in the catch figure for The Netherlands
${ }^{12}$ may include misreported catch from IVaN and discards
${ }^{13}$ These catches (including some local fjord-type spring spawners) are taken by Norway under a separate quota south of $62^{\circ} \mathrm{N}$ and are not included in the Norwegian North Sea catch figure for this area
${ }^{14}$ Figure altered in 2004

Table 2.1.2: Herring caught in the North Sea. Catch in tonnes in Division IVa West. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.

| Country | 199711 | 199811 | 199911 | 200011 | $2001 \quad 11$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 2657 | 4634 | 15359 | 25530 | 17770 |
| Faroe Islands | 1156 | 1246 | 1977 | 205 | 192 |
| France | 362 | 4758 | 6369 | 3210 | 8164 |
| Germany | 4576 | 7753 | 11206 | 5811 | 17753 |
| Netherlands | 6072 | 10917 | 21552 | 15117 | 1750310 |
| Norway | 16869 | 27290 | 31395 | 33164 | 11653 |
| Sweden | 1617 | 315 | 859 | 1479 | - |
| Poland |  |  |  |  | 1418 |
| Russia | 1619 | 452 | - | - | - |
| UK (England) | 49 | 4306 | 7999 | 8859 | 12283 |
| UK (Scotland) | 17121 | 29462 | 28537 | 29055 | 25105 |
| UK (N. Ireland) | - | 1015 | - | 996 | 1018 |
| Unallocated landings | 40662 s,8 | 560588 | 254698 | 443348 | 247258 |
| Misreporting from VIa North |  |  |  |  |  |
| Total Landings | 92760 | 148206 | 150722 | 167760 | 137584 |
| Discards |  |  |  |  |  |
| Total catch | 92760 | 148206 | 150722 | 167760 | 137584 |


| Country | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Denmark 7 | 26422 | 48358 | 48128 | 80990 | 60462 |
| Faroe Islands | - | 95 | - |  | 580 |
| France | 10522 | 11237 | 10941 | 13474 | 18453 |
| Germany | 15189 | 25796 | 17559 | 22278 | 18605 |
| Netherlands | 18289 | 25045 | 43876 | 36619 | 39209 |
| Norway | 10836 | 34443 | 36119 | 66232 | 38363 |
| Poland | - | - | - | 458 | - |
| Sweden | 2397 | 2647 | 2178 | 8261 | 4957 |
| Russia | - | - | - | 99 | - |
| UK (England) | 10142 | 12030 | 13480 | 15523 | 12031 |
| UK (Scotland) | 30014 | 39970 | 43490 | 71941 | 47368 |
| UK (N. Ireland) | 944 | 2010 | 2656 | 2912 | 3531 |
| Unallocated landings | 14201 | 8 | 14115 | 8 | 28631 |
| Misreporting from VIa North |  |  | 39324 | 8 | 10981 |
| Total Landings | 138956 | 215746 | 247058 | 358111 | 253048 |
| Discards | 17093 | 4125 | 15794 | 10861 | 1492 |
| Total catch | $\mathbf{1 5 6 0 4 9}$ | $\mathbf{2 1 9 8 7 1}$ | $\mathbf{2 6 2 8 5 2}$ | $\mathbf{3 6 8 9 7 2}$ | $\mathbf{2 5 4 5 4 0}$ |

[^0]Table 2.1.3: Herring caught in the North Sea. Catch in tonnes in Division IVa East. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.

| Country | $\mathbf{1 9 9 7}$ | $\mathbf{7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{7}$ | $\mathbf{1 9 9 9}$ | $\mathbf{7}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Denmark 5 | 22862 | 25750 | 18259 | 11300 | $\mathbf{7}$ | $\mathbf{2 0 0 1} 7$ |
| Faroe Islands | - | - | - | 710 | 890 |  |
| France | 3 | - | 115 | - | - |  |
| Germany | - | - | - | 29 | - |  |
| Netherlands | 756 | 301 | - | 38 | - |  |
| Norway 2 | 20975 | 43646 | 39977 | 38655 | 56904 |  |
| Sweden | 422 | 1189 | 772 | 1177 | 517 |  |
| Unallocated landings | -756 | 4 | -292 | 4 | - | 338 |
| Total landings | 44262 | 70594 | 59123 | 52247 | 76777 |  |
| Discards | - | - | - | - | - |  |
| Total catch | $\mathbf{4 4 2 6 2}$ | $\mathbf{7 0 5 9 4}$ | $\mathbf{5 9 1 2 3}$ | 52247 | $\mathbf{7 6 7 7 7}$ |  |
| Norw. Spring Spawners 6 | 54728 | 29220 | 32106 | 25678 | 7108 |  |


| Country | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Denmark 5 | 17846 | 7401 | 16278 | 5761 | 8614 |
| Faroe Islands | 1365 | 359 | - | 738 | 975 |
| France | - | - | - |  | - |
| Germany | 81 | 54 | 888 |  | 34 |
| Netherlands | - | - | - |  | - |
| Norway 2 | 63482 | 62306 | 100443 | 89925 | 90065 |
| UK (Scotland) | - | - | - | - | 83 |
| Sweden | 568 | 1529 | 1720 | 3510 | 2857 |
| Unallocated landings | 5961 | 11991 | 0 | 0 | 0 |
| Total landings | 89303 | 83640 | 119329 | 99934 | 102628 |
| Discards | - | - | - | - | - |
| Total catch | $\mathbf{8 9 3 0 3}$ | $\mathbf{8 3 6 4 0}$ | $\mathbf{1 1 9 3 2 9}$ | $\mathbf{9 9 9 3 4}$ | $\mathbf{1 0 2 6 2 8}$ |
| Norw. Spring Spawners 6 | 4069 | 979 | 452 | 417 | 626 |

[^1]Table 2.1.4: Herring caught in the North Sea. Catch in tonnes in Division IVb. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.


| Country | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Belgium | - | - | - | - | - |
| Denmark 4 | 26387 | 22574 | 33857 | 41423 | 32277 |
| Faroe Islands | 48 | 173 | 402 | - | 200 |
| France | 4214 | 7918 | 10592 | 10205 | 17385 |
| Germany | 7577 | 12116 | 13823 | 14381 | 14222 |
| Netherlands | 13154 | 19115 | 23649 | 10038 | 13363 |
| Norway | 656 | 15732 | 1076 | 645 | 6933 |
| Sweden | 453 | 605 | 1794 | 1694 | 2715 |
| UK (England) | 317 | 2632 | 2864 | 3869 | 4924 |
| UK (Scotland) | 289 | 322 | 1841 | 1286 | 977 |
| Unallocated landings | 4052 | -2401 | 8300 | 10233 | 2364 |
| Total landings | 57147 | 78786 | 98198 | 93774 | 95360 |
| Discards 2 |  |  | 1265 | 1963 |  |
| Total catch | $\mathbf{5 7 1 4 7}$ | $\mathbf{7 8 7 8 6}$ | $\mathbf{9 9 4 6 3}$ | $\mathbf{9 5 7 3 7}$ | $\mathbf{9 5 3 6 0}$ |

[^2]Table 2.1.5: Herring caught in the North Sea. Catch in tonnes in Division IVc and VIId. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.

| Country | 1997 | 9 | 1998 | 9 | 1999 | 9 | 2000 | 9 | 20019 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 1 |  | - |  | 1 |  | 1 |  | - |
| Denmark | 1247 |  | 1873 |  | 1439 |  | 468 |  | 583 |
| France | 8091 |  | 7081 |  | 12844 |  | 6879 |  | 8750 |
| Germany | 1349 |  | 916 |  | 2029 |  | 2029 |  | 3686 |
| Netherlands | 14181 |  | 11247 |  | 10572 |  | 12348 |  | 9630 |
| UK (England) | 1388 |  | 1562 |  | 1794 |  | 1651 |  | 1485 |
| UK (Scotland) | - |  | - |  | - |  | - |  | - |
| Unallocated landings | 27241 | 4 | 26701 | 4 | 21652 | 4 | 26822 | 4 | 25522 4 |
| Total landings Discards 3 | 53498 |  | 49380 |  | 50331 |  | 50198 |  | 49656 |
| Total catch | 53498 |  | 49380 |  | 50331 |  | 50198 |  | 49656 |
| Coastal spring spawners included above 2 | 143 |  | 88 |  | 88 |  | 76 |  | 14711 |


| Country | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Belgium | 23 | 5 | 8 | 6 | 3 |
| Denmark | 170 | 273 | 774 | 206 | 969 |
| Faroe Islands | - | - | - | - | 30 |
| France | 10686 | 12389 | 12988 | 15150 | 13637 |
| Germany | 4366 | 5987 | 9588 | 9896 | 7553 |
| Netherlands | 23814 | 36948 | 28637 | 34874 | 23743 |
| UK (England) | 3298 | 3977 | 4511 | 5919 | 5243 |
| UK (Scotland) | 623 | - | - | - | - |
| Unallocated landings | 7338 | 8170 | 11967 | 8231 | 5419 |
| Total landings | 50318 | 67749 | 68473 | 74282 | 56597 |
| Discards 3 | - | - | - | - | - |
| Total catch | $\mathbf{5 0 3 1 8}$ | $\mathbf{6 7 7 4 9}$ | $\mathbf{6 8 4 7 3}$ | $\mathbf{7 4 2 8 2}$ | $\mathbf{5 6 5 9 7}$ |
| Coastal spring spawners | 60 | 84 | 62 | 74 | 65 |
| included above 2 |  |  |  |  |  |

[^3]Table 2.1.6 ("The Wonderful Table"): HERRING in Sub-area IV, Division VIId and Division IIIa. Figures in thousand tonnes.

| $\overline{\text { Year }}$ | 1989 | 1990 |  | 1991 |  | 1992 | 1993 | 1994 | 199518 | $8 \quad 199618$ | 199718 | 199818 | 199918 | 200018 | 200118 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sub-Area IV and Division VIId: TAC (IV and VIId) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Recommended Divisions IVa, b 1 | 484 | 373, 332 |  | 363 | 6 | 352 | 2907 | 2967 | 38911 | 156 | 159 | 254 | 265 | 265 | - 22 | - 22 | - 22 | - 22 | - 22 | - 22 |
| Recommended Divisions IVc, VIId | 30 | 30 |  | 50-60 | 6 | 54 | 50 | 50 | 50 | 14 | 14 | 14 | - 14 | 14 | -14 | - 14 | - 14 | 14 | 14 | - 14 |
| Expected catch of spring spawners |  |  |  |  |  | 10 | 8 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Agreed Divisions IVa, b 2 | 484 | 385 |  | 370 | 6 | 380 | 380 | 390 | 390 | 263;131 13 | 134 | 229 | 240 | 240 | 240 | 223 | 340.5 | 393.9 | 460.7 | 404.7 |
| Agreed Div. IVc, VIId | 30 | 30 |  | 50 | 6 | 50 | 50 | 50 | 50 | 50; 2513 | 25 | 25 | 25 | 25 | 25 | 42.7 | 59.5 | 66.1 | 74.3 | 50.0 |
| Bycatch ceiling in the small mesh fishery |  |  |  |  |  |  |  |  |  |  | 24 | 22 | 30 | 36 | 36 | 36 | 52.0 | 38.0 | 50.0 | 42.5 |
| CATCH (IV and VIId) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| National landings Divisions IVa, 3 | 639 | 499 |  | 495 |  | 481 | 463 | 421 | 465 | 183 | 149 | 245 | 261 | 261 | 272 | 261 | 354.5 | 427.7 | 502.3 | 439.2 |
| Unallocated landings Divisions IVa, b | -2 | 14 |  | 30 |  | 14 | -1 | 6 | -15 | -5 | 36 | 44 | 22 | 35 | 2 | 24 | 23.7 | 36.9 | 49.6 | 13.3 |
| Discard/slipping Divisions IVa, 4 | 3 | 4 |  | 2 |  | 3 | 1 | 1 |  | - |  |  | - |  | - | 17 | 4.1 | 17.1 | 12.8 | 1.5 |
| Total catch Divisions IVa,b 5 | 638 | 516 |  | 527 |  | 498 | 463 | 428 | 450 | 178 | 185 | 289 | 283 | 296 | 273 | 303 | 382.3 | 481.6 | 564.6 | 454.0 |
| National landings Divisions IVc, VIId 3 | 30 | 24 |  | 42 |  | 37 | 3221 | 42 | 45 | 24 | 26 | 23 | 29 | 23 | 24 | 43 | 59.5 | 56.5 | 66.1 | 51.2 |
| Unallocated landings Divisions IVc,VIId | 48 | 32 |  | 16 |  | 35 | 43 | 30 | 22 | 31 | 27 | 27 | 22 | 27 | 26 | 7 | 8.2 | 12.0 | 8.2 | 5.4 |
| Discard/slipping Divisions IVc, VIId 4 | 1 | 5 |  | 3 |  | , | 2 | 2 | - | - | - | - | - | - | - | 0 | - | - | - | - |
| Total catch Divisions IVc, VIId | 79 | 61 |  | 61 |  | 74 | 7721 | 74 | 67 | 55 | 53 | 49 | 50 | 50 | 50 | 50 | 67.7 | 68.5 | 74.3 | 56.6 |
| Total catch IV and VIId as used by ACFM 5 | 717 | 578 |  | 588 |  | 572 | 54021 | 498 | 516 | 233 | 238 | 338 | 333 | 346 | 323 | 353 | 450.0 | 550.1 | 638.9 | 510.6 |
| CATCH BY FLEET/STOCK (IV and VIId) 10 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| North Sea autumn spawners directed fisheries (Fleet A) | N.a. | N.a. |  | 446 |  | 441 | 438 | 447 | 439 | 195 | 225 | 316 | 313 | 322 | 296 | 323 | 434.9 | 529.5 | 610.0 | 487.1 |
| North Sea autumn spawners industrial (Fleet B) | N.a. | N.a. |  | 134 |  | 124 | 101 | 38 | 67 | 38 | 13 | 14 | 15 | 18 | 20 | 22 | 12.3 | 13.6 | 21.8 | 11.9 |
| North Sea autumn spawners in IV and VIId total | 696 | 569 |  | 580 |  | 564 | 539 | 485 | 506 | 233 | 237 | 330 | 329 | 339 | 317 | 346 | 447.2 | 543.0 | 631.9 | 499.0 |
| Baltic-III-type spring spawners in IV | 20 | 8 |  | 8 |  | 8 | 9 | 13 | 10 | 1 | 1 | 8 | 5 | 7 | 6 | 7 | 2.8 | 7.1 | 7.0 | 11.0 |
| Coastal-type spring spawners | 2.3 | 1.1 |  | 0.3 |  | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.1 | 0.1 | 0.1 | 1.2 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| Norw. Spring Spawners caught under a separate quota in IV 20 | N.a. | , |  | 5 |  | 5 | 9 | 6 | 10 | 30 | 55 | 29 | 32 | 26 | 7 | 4 | 1.0 | 0.5 | 0.4 | 0.6 |
| Division IIIa: TAC (IIIa) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Predicted catch of autumn spawners |  |  |  | 96 |  | 153 | 102 | 77 | 98 | 48 | 35 | 58 | 43 | 53 | - 22 | - | -22 | -2 | - 22 | - 22 |
| Recommended spring spawners | 84 | 67 |  | 91 |  | 90 | 93-113 | 9 | 12 | - -12 | - 15 | - 15 | - 15 | - 15 | - 15 | 15 | 15 | 22 | 22 | - 22 |
| Recommended mixed clupeoids | 80 | 60 |  | 0 |  | 0 | 0 | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Agreed herring TAC | 138 | 120 |  | 104.5 |  | 124 | 165 | 148 | 140 | 120 | 80 | 80 | 80 | 80 | 80 | 80 | 80.0 | 70.0 | 96.0 | 81.6 |
| Agreed mixed clupeoid TAC | 80 | 65 |  | 50 |  | 50 | 45 | 43 | 43 | 43 |  |  |  |  |  |  |  |  |  |  |
| Bycatch ceiling in the small mesh fishery |  |  |  |  |  |  |  |  |  |  | 20 | 17 | 19 | 21 | 21 | 21 | 21.0 | 21.0 | 24.2 | 20.5 |
| CATCH (IIIa) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| National landings | 192 | 202 |  | 188 |  | 227 | 214 | 168 | 157 | 115 | 83 | 120 | 86 | 108 | 90 | 79 | 76.0 | 61.1 | 90.8 | 88.9 |
| Catch as used by ACFM | 162 | 195 |  | 191 |  | 227 | 214 | 168 | 140 | 105 | 74 | 108 | 79 | 99 | 82 | 73 | 68.1 | 52.7 | 69.6 | 51.2 |
| CATCH BY FLEET/STOCK (IIIa) 10 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Autumn spawners human consumption (Fleet C) | N.a. | N.a. |  | 26 |  | 47 | 44 | 42 | 38 | 24 | 21 | 59 | 2817 | 36 | 34 | 17 | 24.1 | 13.4 | 22.9 | 11.6 |
| Autumn spawners mixed clupeoid (Fleet D) 19 | N.a. | N.a. |  | 13 |  | 23 | 25 | 12 | 6 | 9 | 4 | 6 | $8{ }^{17}$ | 13 | 12 | 9 | 8.4 | 10.8 | 9.0 | 3.4 |
| Autumn spawners other industrial landings (Fleet E) | N.a. | N.a. |  | 38 |  | 82 | 63 | 32 | 29 | 8 | 2 |  |  |  |  |  |  |  |  |  |
| Autumn spawners in IIIa total | 91 | 77 | 8 | 77 |  | 152 | 132 | 86 | 73 | 43 | 27 | 61 | 3417 | 49 | 46 | 26 | 32.5 | 24.2 | 31.9 | 15.0 |
| Spring spawners human consumption (Fleet C) | N.a. | N.a. |  | 68 |  | 53 | 68 | 59 | 44 | 58 | 43 | 40 | 4017 | 45 | 33 | 38 | 31.6 | 16.8 | 32.5 | 30.2 |
| Spring spawners mixed clupeoid (Fleet D) 19 | N.a. | N.a. |  | 5 |  | 2 | 1 | 1 | 2 | 4 | 3 | 3 | 317 | 5 | 3 | 9 | 4.0 | 11.2 | 5.1 | 5.9 |
| Spring spawners other industrial landings (Fleet E) | N.a. | N.a. |  | 40 |  | 20 | 12 | 24 | 21 | 2 | 1 |  |  |  |  |  |  |  |  |  |
| Spring spawners in IIIa total | 71 | 118 |  | 113 |  | 75 | 81 | 84 | 67 | 64 | 47 | 43 | 4317 | 50 | 36 | 47 | 35.6 | 28.0 | 37.6 | 36.1 |
| North Sea autumn spawners Total as used by ACFM | 787 | 646 |  | 657 |  | 716 | 671 | 571 | 579 | 275 | 264 | 392 | 363 | 388 | 363 | 372 | 479.7 | 567.2 | 663.8 | 514.6 |

1 Includes catches in directed fishery and catches of 1-ringers in small mesh fishery up to 1992.2 IVa,b and EC zone of IIa. 3 Provided by Working Group members. 4 Incomplete, only some countries providing discard information. Discards might also be included in un. 5 Includes spring spawners not included in assessment. 6 Revised during 1991. 7 Based on F=0.3 in directed fishery only; TAC advised for IVC, VIId subtracted. 8 Estimated. 9 130-180 for spring spawners in all areas. 10 Based on sum-ofproducts (number x mean weight at age). 11 Status quo F catch for fleet A. 12 The catch should not exceed recent catch levels. 13 During the middle of 1996 revised to $50 \%$ of its original agreed TAC. 14 Included in IVa,b. 15 Managed in accordance with autumn spawners. 17 Figure altered in 2001 and again in 2004.18 Data for $1995-2001$ were verified and amended where necessary by SG REDNOSE in 2003 . 19 Fleet D and E are merged from 1999 onwards. 20 These catches (including local fjord-typ
Spring Spawners) are taken by Norway under a separate quota south of $62^{\circ} \mathrm{N}$ and are not included in the Norwegian North Sea catch figure for this area. 21 figure altered in 2003 to account for earlier summarizing errors. 22 See catch option tables for different fleets.Shaded cells for the catch by fleet in Division IIIa indicate persisting inconsistencies which have to be resolved intersessionally.

Table 2.2.1: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea in 2006. Catch in numbers (millions) at age (CANUM), by quarter and division.


Quarters: 1-4

| 0 | 35.1 | 0.0 | 0.0 | 0.0 | 5.1 | 837.2 | 0.3 | 0.9 | 842.4 | 1.2 | 878.6 | 843.5 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 150.1 | 0.4 | 0.1 | 0.3 | 4.1 | 37.8 | 21.3 | 8.6 | 42.1 | 29.9 | 222.1 | 72.1 |
| 2 | 50.2 | 46.7 | 3.5 | 43.2 | 115.5 | 104.8 | 14.1 | 73.3 | 263.5 | 87.4 | 401.1 | 354.4 |
| 3 | 10.2 | 62.4 | 8.8 | 53.7 | 127.7 | 60.4 | 4.1 | 54.5 | 241.8 | 58.6 | 310.6 | 309.2 |
| 4 | 3.3 | 126.7 | 14.0 | 112.7 | 226.2 | 86.7 | 1.8 | 34.4 | 425.6 | 36.1 | 465.0 | 475.3 |
| 5 | 3.3 | 195.0 | 22.4 | 172.6 | 486.9 | 182.8 | 3.4 | 149.4 | 842.4 | 152.8 | 998.5 | 1016.8 |
| 6 | 0.6 | 51.9 | 5.1 | 46.8 | 150.0 | 38.8 | 0.4 | 16.4 | 235.6 | 16.9 | 253.1 | 256.7 |
| 7 | 0.4 | 63.6 | 5.3 | 58.3 | 139.0 | 35.6 | 0.1 | 15.1 | 232.9 | 15.2 | 248.5 | 251.9 |
| 7 | 0.2 | 19.7 | 2.1 | 17.6 | 30.4 | 9.5 | 0.1 | 5.5 | 57.5 | 5.5 | 63.2 | 65.0 |
| 9 | 0.0 | 12.9 | 1.0 | 11.9 | 20.5 | 5.8 | 0.1 | 5.5 | 38.2 | 5.6 | 43.8 | 44.3 |
| Sum | 253.3 | 579.2 | 62.2 | 517.1 | 1305.5 | 1399.5 | 45.7 | 363.4 | 3222.1 | 409.1 | 3884.6 | 3689.3 |

Quarter: 1

| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 0.0 | 0.9 | 0.3 | 0.9 | $\mathbf{1 . 2}$ | $\mathbf{1 . 2}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 49.4 | 0.4 | 0.1 | 0.2 | 0.4 | 9.6 | 21.1 | 8.6 | 10.2 | 29.7 | $\mathbf{8 9 . 2}$ | $\mathbf{4 0 . 0}$ |
| 2 | 26.5 | 2.7 | 0.5 | 2.2 | 4.2 | 47.4 | 6.7 | 1.8 | 53.8 | 8.5 | $\mathbf{8 8 . 7}$ | $\mathbf{6 2 . 7}$ |
| 3 | 6.1 | 5.5 | 1.0 | 4.5 | 19.9 | 4.4 | 0.7 | 18.9 | 28.9 | 19.6 | $\mathbf{5 4 . 6}$ | $\mathbf{4 9 . 5}$ |
| 4 | 1.5 | 12.7 | 0.2 | 12.5 | 40.1 | 0.5 | 0.3 | 11.3 | 53.1 | 11.6 | $\mathbf{6 6 . 2}$ | $\mathbf{6 4 . 8}$ |
| 5 | 2.3 | 18.4 | 0.2 | 18.2 | 53.1 | 1.5 | 1.6 | 58.2 | 72.8 | 59.8 | $\mathbf{1 3 4 . 9}$ | $\mathbf{1 3 2 . 8}$ |
| 6 | 0.3 | 1.2 | 0.0 | 1.2 | 13.3 | 0.1 | 0.2 | 5.4 | 14.5 | 5.6 | $\mathbf{2 0 . 3}$ | $\mathbf{2 0 . 1}$ |
| 7 | 0.1 | 3.5 | 0.0 | 3.5 | 10.8 | 0.1 | 0.1 | 3.2 | 14.4 | 3.2 | $\mathbf{1 7 . 7}$ | $\mathbf{1 7 . 6}$ |
| 8 | 0.1 | 2.3 | 0.0 | 2.3 | 1.5 | 0.0 | 0.1 | 1.8 | 3.8 | 1.9 | $\mathbf{5 . 8}$ | $\mathbf{5 . 7}$ |
| $9+$ | 0.0 | 0.1 | 0.0 | 0.1 | 0.8 | 0.0 | 0.1 | 2.7 | 1.0 | 2.8 | $\mathbf{3 . 7}$ | $\mathbf{3 . 7}$ |
| Sum | $\mathbf{8 6 . 3}$ | $\mathbf{4 6 . 8}$ | $\mathbf{2 . 0}$ | $\mathbf{4 4 . 7}$ | $\mathbf{1 4 4 . 1}$ | $\mathbf{6 3 . 8}$ | $\mathbf{3 0 . 8}$ | $\mathbf{1 1 2 . 7}$ | $\mathbf{2 5 2 . 6}$ | $\mathbf{1 4 3 . 5}$ | $\mathbf{4 8 2 . 4}$ | $\mathbf{3 9 8 . 2}$ |

## Quarter: 2

| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | $\mathbf{0 . 0}$ | $\mathbf{0 . 0}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 10.0 | 0.0 | 0.0 | 0.0 | 0.0 | 5.9 | 0.2 | 0.0 | 5.9 | 0.2 | $\mathbf{1 6 . 0}$ | $\mathbf{6 . 1}$ |
| 2 | 11.3 | 38.9 | 3.0 | 35.9 | 25.3 | 8.1 | 0.0 | 0.0 | 69.2 | 0.0 | $\mathbf{8 0 . 6}$ | $\mathbf{7 2 . 3}$ |
| 3 | 0.0 | 47.0 | 5.2 | 41.7 | 32.1 | 12.5 | 0.0 | 0.0 | 86.3 | 0.0 | $\mathbf{8 6 . 4}$ | $\mathbf{9 1 . 6}$ |
| 4 | 0.0 | 98.4 | 12.0 | 86.4 | 47.3 | 10.1 | 0.0 | 0.0 | 143.9 | 0.0 | $\mathbf{1 4 3 . 9}$ | $\mathbf{1 5 5 . 8}$ |
| 5 | 0.0 | 136.1 | 16.5 | 119.6 | 119.0 | 23.5 | 0.0 | 0.0 | 262.0 | 0.0 | $\mathbf{2 6 2 . 1}$ | $\mathbf{2 7 8 . 6}$ |
| 6 | 0.0 | 24.5 | 3.0 | 21.5 | 21.8 | 3.4 | 0.0 | 0.0 | 46.7 | 0.0 | $\mathbf{4 6 . 7}$ | $\mathbf{4 9 . 7}$ |
| 7 | 0.0 | 19.3 | 2.3 | 16.9 | 22.3 | 4.3 | 0.0 | 0.0 | 43.5 | 0.0 | $\mathbf{4 3 . 5}$ | $\mathbf{4 5 . 8}$ |
| 8 | 0.0 | 10.3 | 1.3 | 9.1 | 5.9 | 2.3 | 0.0 | 0.0 | 17.3 | 0.0 | $\mathbf{1 7 . 3}$ | $\mathbf{1 8 . 6}$ |
| $9+$ | 0.0 | 3.0 | 0.4 | 2.6 | 4.4 | 0.7 | 0.0 | 0.0 | 7.8 | 0.0 | $\mathbf{7 . 8}$ | $\mathbf{8 . 1}$ |
| Sum | $\mathbf{2 1 . 4}$ | $\mathbf{3 7 7 . 5}$ | $\mathbf{4 3 . 6}$ | $\mathbf{3 3 3 . 8}$ | $\mathbf{2 7 8 . 1}$ | $\mathbf{7 0 . 7}$ | $\mathbf{0 . 3}$ | $\mathbf{0 . 0}$ | $\mathbf{6 8 2 . 7}$ | $\mathbf{0 . 3}$ | $\mathbf{7 0 4 . 4}$ | $\mathbf{7 2 6 . 7}$ |

Quarter: 3

| 0 | 27.0 | 0.0 | 0.0 | 0.0 | 5.1 | 651.7 | 0.0 | 0.0 | 656.8 | 0.0 | $\mathbf{6 8 3 . 7}$ | $\mathbf{6 5 6 . 8}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 40.8 | 0.0 | 0.0 | 0.0 | 1.9 | 17.3 | 0.0 | 0.0 | 19.2 | 0.0 | $\mathbf{6 0 . 0}$ | $\mathbf{1 9 . 2}$ |
| 2 | 10.2 | 4.9 | 0.1 | 4.9 | 72.0 | 28.3 | 0.0 | 0.0 | 105.2 | 0.0 | $\mathbf{1 1 5 . 4}$ | $\mathbf{1 0 5 . 3}$ |
| 3 | 2.5 | 8.7 | 2.4 | 6.3 | 67.8 | 25.8 | 0.0 | 0.0 | 99.8 | 0.0 | $\mathbf{1 0 2 . 4}$ | $\mathbf{1 0 2 . 3}$ |
| 4 | 0.8 | 5.7 | 1.3 | 4.4 | 111.6 | 63.0 | 0.0 | 0.0 | 179.0 | 0.0 | $\mathbf{1 7 9 . 9}$ | $\mathbf{1 8 0 . 4}$ |
| 5 | 0.3 | 19.3 | 4.5 | 14.8 | 274.7 | 88.6 | 0.0 | 0.0 | 378.1 | 0.0 | $\mathbf{3 7 8 . 5}$ | $\mathbf{3 8 2 . 7}$ |
| 6 | 0.2 | 4.1 | 0.9 | 3.1 | 104.0 | 20.3 | 0.0 | 0.0 | 127.4 | 0.0 | $\mathbf{1 2 7 . 6}$ | $\mathbf{1 2 8 . 4}$ |
| 7 | 0.1 | 4.1 | 1.0 | 3.1 | 98.5 | 22.5 | 0.0 | 0.0 | 124.1 | 0.0 | $\mathbf{1 2 4 . 2}$ | $\mathbf{1 2 5 . 1}$ |
| 8 | 0.0 | 2.5 | 0.6 | 1.9 | 20.0 | 3.4 | 0.0 | 0.0 | 25.3 | 0.0 | $\mathbf{2 5 . 4}$ | $\mathbf{2 5 . 9}$ |
| $9+$ | 0.0 | 0.5 | 0.1 | 0.4 | 14.8 | 5.0 | 0.0 | 0.0 | 20.2 | 0.0 | $\mathbf{2 0 . 2}$ | $\mathbf{2 0 . 3}$ |
| Sum | $\mathbf{8 2 . 0}$ | $\mathbf{4 9 . 8}$ | $\mathbf{1 0 . 9}$ | $\mathbf{3 8 . 9}$ | $\mathbf{7 7 0 . 4}$ | $\mathbf{9 2 5 . 9}$ | $\mathbf{0 . 0}$ | $\mathbf{0 . 1}$ | $\mathbf{1 7 3 5 . 2}$ | $\mathbf{0 . 1}$ | $\mathbf{1 8 1 7 . 4}$ | $\mathbf{1 7 4 6 . 2}$ |


| Quarter: 4 |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 8.1 | 0.0 | 0.0 | 0.0 | 0.0 | 185.3 | 0.3 | 0.0 | 185.3 | 0.3 | $\mathbf{1 9 3 . 7}$ | $\mathbf{1 8 5 . 6}$ |
| 1 | 49.9 | 0.0 | 0.0 | 0.0 | 1.8 | 5.1 | 0.0 | 0.0 | 6.9 | 0.0 | $\mathbf{5 6 . 8}$ | $\mathbf{6 . 9}$ |
| 2 | 2.2 | 0.2 | 0.0 | 0.2 | 14.1 | 21.0 | 7.4 | 71.4 | 35.3 | 78.9 | $\mathbf{1 1 6 . 3}$ | $\mathbf{1 1 4 . 2}$ |
| 3 | 1.5 | 1.3 | 0.2 | 1.1 | 8.0 | 17.7 | 3.4 | 35.5 | 26.8 | 38.9 | $\mathbf{6 7 . 3}$ | $\mathbf{6 5 . 9}$ |
| 4 | 0.8 | 9.9 | 0.5 | 9.4 | 27.1 | 13.1 | 1.4 | 23.1 | 49.7 | 24.5 | $\mathbf{7 5 . 0}$ | $\mathbf{7 4 . 3}$ |
| 5 | 0.7 | 21.1 | 1.1 | 20.0 | 40.2 | 69.3 | 1.8 | 91.1 | 129.4 | 92.9 | $\mathbf{2 2 3 . 0}$ | $\mathbf{2 2 2 . 7}$ |
| 6 | 0.1 | 22.2 | 1.2 | 21.0 | 10.9 | 15.1 | 0.3 | 11.0 | 47.0 | 11.3 | $\mathbf{5 8 . 3}$ | $\mathbf{5 8 . 5}$ |
| 7 | 0.2 | 36.7 | 1.9 | 34.8 | 7.4 | 8.8 | 0.0 | 11.9 | 50.9 | 12.0 | $\mathbf{6 3 . 1}$ | $\mathbf{6 3 . 3}$ |
| 8 | 0.1 | 4.6 | 0.2 | 4.3 | 3.0 | 3.7 | 0.0 | 3.7 | 11.0 | 3.7 | $\mathbf{1 4 . 7}$ | $\mathbf{1 4 . 7}$ |
| $9+$ | 0.0 | 9.2 | 0.5 | 8.8 | 0.6 | 0.0 | 0.0 | 2.8 | 9.3 | $\mathbf{2 . 8}$ | $\mathbf{1 2 . 9}$ | $\mathbf{1 2 . 2}$ |
| Sum | $\mathbf{6 3 . 6}$ | $\mathbf{1 0 5 . 2}$ | $\mathbf{5 . 6}$ | $\mathbf{9 9 . 6}$ | $\mathbf{1 1 2 . 9}$ | $\mathbf{3 3 9 . 1}$ | $\mathbf{1 4 . 6}$ | $\mathbf{2 5 0 . 6}$ | $\mathbf{5 5 1 . 6}$ | $\mathbf{2 6 5 . 2}$ | $\mathbf{8 8 0 . 3}$ | $\mathbf{8 1 8 . 3}$ |

Table 2.2.2: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea in 2006. Mean weight-at-age (kg) in the catch (WECA), by quarter and division.

| WR | $\begin{array}{r} \text { IIIa } \\ \text { NSAS } \end{array}$ | $\begin{array}{r} \text { IVa(E) } \\ \text { all } \end{array}$ | $\begin{aligned} & \text { IVa(E) } \\ & \text { WBSS } \end{aligned}$ | IVa(W) | IVb | IVc | VIId | IVa \& IVb all | $\begin{array}{r} \text { IVc \& } \\ \text { VIId } \end{array}$ | $\begin{aligned} & \text { Total } \\ & \text { NSAS } \end{aligned}$ | Herring caught in the North Sea |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

Quarters: 1-4

|  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0.014 | 0.000 | 0.000 | 0.010 | 0.010 | 0.016 | 0.081 | 0.010 | - | 0.010 | 0.010 |
| 1 | 0.054 | 0.025 | 0.111 | 0.103 | 0.020 | 0.027 | 0.130 | 0.028 | 0.057 | 0.049 | 0.040 |
| 2 | 0.079 | 0.125 | 0.108 | 0.145 | 0.097 | 0.090 | 0.125 | 0.123 | 0.119 | 0.117 | 0.122 |
| 3 | 0.117 | 0.149 | 0.152 | 0.156 | 0.141 | 0.130 | 0.124 | 0.150 | 0.125 | 0.144 | 0.145 |
| 4 | 0.140 | 0.164 | 0.171 | 0.180 | 0.172 | 0.151 | 0.153 | 0.174 | 0.153 | 0.172 | 0.172 |
| 5 | 0.186 | 0.175 | 0.178 | 0.193 | 0.183 | 0.150 | 0.152 | 0.187 | 0.152 | 0.181 | 0.181 |
| 6 | 0.191 | 0.214 | 0.191 | 0.230 | 0.202 | 0.195 | 0.177 | 0.222 | 0.178 | 0.220 | 0.220 |
| 7 | 0.216 | 0.224 | 0.189 | 0.251 | 0.220 | 0.170 | 0.205 | 0.239 | 0.205 | 0.237 | 0.237 |
| 8 | 0.207 | 0.229 | 0.214 | 0.247 | 0.232 | 0.195 | 0.209 | 0.238 | 0.209 | 0.235 | 0.235 |
| $9+$ | 0.000 | 0.254 | 0.201 | 0.286 | 0.239 | 0.216 | 0.220 | 0.269 | 0.219 | 0.262 | 0.262 |

Quarter: 1

| Quarter: 1 |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0.000 | 0.020 | 0.020 | 0.020 | 0.020 | 0.081 | 0.081 | - | - | $\mathbf{0 . 0 6 7}$ | $\mathbf{0 . 0 6 7}$ |
| 1 | 0.020 | 0.024 | 0.024 | 0.024 | 0.024 | 0.027 | 0.130 | 0.024 | 0.057 | $\mathbf{0 . 0 3 3}$ | $\mathbf{0 . 0 4 8}$ |
| 2 | 0.070 | 0.107 | 0.107 | 0.097 | 0.070 | 0.048 | 0.117 | 0.074 | 0.062 | $\mathbf{0 . 0 7 2}$ | $\mathbf{0 . 0 7 2}$ |
| 3 | 0.107 | 0.127 | 0.127 | 0.126 | 0.107 | 0.095 | 0.092 | 0.123 | 0.092 | $\mathbf{0 . 1 1 0}$ | $\mathbf{0 . 1 1 1}$ |
| 4 | 0.129 | 0.140 | 0.140 | 0.133 | 0.141 | 0.109 | 0.109 | 0.134 | 0.109 | $\mathbf{0 . 1 3 0}$ | $\mathbf{0 . 1 3 0}$ |
| 5 | 0.187 | 0.151 | 0.151 | 0.147 | 0.145 | 0.124 | 0.124 | 0.148 | 0.124 | $\mathbf{0 . 1 3 8}$ | $\mathbf{0 . 1 3 7}$ |
| 6 | 0.195 | 0.193 | 0.193 | 0.165 | 0.171 | 0.153 | 0.153 | 0.167 | 0.153 | $\mathbf{0 . 1 6 3}$ | $\mathbf{0 . 1 6 3}$ |
| 7 | 0.220 | 0.164 | 0.164 | 0.172 | 0.177 | 0.166 | 0.166 | 0.170 | 0.166 | $\mathbf{0 . 1 6 9}$ | $\mathbf{0 . 1 6 9}$ |
| 8 | 0.209 | 0.189 | 0.189 | 0.186 | 0.192 | 0.194 | 0.194 | 0.188 | 0.194 | $\mathbf{0 . 1 9 0}$ | $\mathbf{0 . 1 9 0}$ |
| $9+$ | 0.000 | 0.235 | 0.235 | 0.219 | 0.213 | 0.216 | 0.216 | 0.221 | - | $\mathbf{0 . 2 1 7}$ | $\mathbf{0 . 2 1 7}$ |

Quarter: 2

| 0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.081 | 0.081 | - | - | $\mathbf{0 . 0 8 1}$ | $\mathbf{0 . 0 8 1}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.045 | 0.000 | 0.000 | 0.000 | 0.031 | 0.029 | 0.130 | 0.031 | 0.030 | $\mathbf{0 . 0 4 0}$ | $\mathbf{0 . 0 3 1}$ |
| 2 | 0.075 | 0.126 | 0.126 | 0.134 | 0.100 | 0.057 | 0.117 | 0.126 | 0.057 | $\mathbf{0 . 1 1 9}$ | $\mathbf{0 . 1 2 6}$ |
| 3 | 0.103 | 0.141 | 0.141 | 0.147 | 0.122 | 0.113 | 0.092 | 0.140 | 0.112 | $\mathbf{0 . 1 4 0}$ | $\mathbf{0 . 1 4 0}$ |
| 4 | 0.135 | 0.160 | 0.160 | 0.181 | 0.144 | 0.124 | 0.109 | 0.165 | 0.121 | $\mathbf{0 . 1 6 6}$ | $\mathbf{0 . 1 6 5}$ |
| 5 | 0.144 | 0.168 | 0.168 | 0.190 | 0.152 | 0.133 | 0.124 | 0.176 | 0.131 | $\mathbf{0 . 1 7 7}$ | $\mathbf{0 . 1 7 6}$ |
| 6 | 0.000 | 0.191 | 0.191 | 0.211 | 0.178 | 0.157 | 0.153 | 0.199 | 0.156 | $\mathbf{0 . 1 9 9}$ | $\mathbf{0 . 1 9 9}$ |
| 7 | 0.000 | 0.206 | 0.206 | 0.241 | 0.188 | 0.161 | 0.166 | 0.222 | 0.162 | $\mathbf{0 . 2 2 3}$ | $\mathbf{0 . 2 2 2}$ |
| 8 | 0.171 | 0.218 | 0.218 | 0.224 | 0.199 | 0.178 | 0.194 | 0.218 | 0.182 | $\mathbf{0 . 2 1 8}$ | $\mathbf{0 . 2 1 8}$ |
| $9+$ | 0.000 | 0.235 | 0.235 | 0.255 | 0.213 | 0.216 | 0.216 | 0.244 | - | $\mathbf{0 . 2 4 4}$ | $\mathbf{0 . 2 4 4}$ |

Quarter: 3

| 0 | 0.013 | 0.000 | 0.000 | 0.010 | 0.010 | 0.000 | 0.000 | 0.010 | - | $\mathbf{0 . 0 1 0}$ | $\mathbf{0 . 0 1 0}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.078 | 0.111 | 0.111 | 0.100 | 0.016 | 0.000 | 0.000 | 0.024 | \#DIV/O! | $\mathbf{0 . 0 6 1}$ | $\mathbf{0 . 0 2 4}$ |
| 2 | 0.103 | 0.149 | 0.149 | 0.152 | 0.118 | 0.000 | 0.125 | 0.143 | 0.125 | $\mathbf{0 . 1 3 9}$ | $\mathbf{0 . 1 4 3}$ |
| 3 | 0.129 | 0.174 | 0.174 | 0.169 | 0.150 | 0.000 | 0.142 | 0.164 | 0.142 | $\mathbf{0 . 1 6 3}$ | $\mathbf{0 . 1 6 4}$ |
| 4 | 0.146 | 0.191 | 0.191 | 0.201 | 0.177 | 0.000 | 0.170 | 0.192 | 0.170 | $\mathbf{0 . 1 9 2}$ | $\mathbf{0 . 1 9 2}$ |
| 5 | 0.163 | 0.193 | 0.193 | 0.205 | 0.205 | 0.000 | 0.167 | 0.205 | 0.167 | $\mathbf{0 . 2 0 5}$ | $\mathbf{0 . 2 0 5}$ |
| 6 | 0.175 | 0.259 | 0.259 | 0.246 | 0.222 | 0.000 | 0.188 | 0.242 | 0.188 | $\mathbf{0 . 2 4 2}$ | $\mathbf{0 . 2 4 2}$ |
| 7 | 0.188 | 0.246 | 0.246 | 0.263 | 0.236 | 0.000 | 0.214 | 0.257 | 0.214 | $\mathbf{0 . 2 5 7}$ | $\mathbf{0 . 2 5 7}$ |
| 8 | 0.204 | 0.244 | 0.244 | 0.262 | 0.255 | 0.000 | 0.219 | 0.259 | 0.219 | $\mathbf{0 . 2 6 0}$ | $\mathbf{0 . 2 5 9}$ |
| $9+$ | 0.000 | 0.310 | 0.310 | 0.299 | 0.243 | 0.000 | 0.224 | 0.285 | 0.224 | $\mathbf{0 . 2 8 5}$ | $\mathbf{0 . 2 8 5}$ |

Quarter: 4

|  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0.018 | 0.000 | 0.000 | 0.000 | 0.010 | 0.010 | 0.000 | 0.010 | - | $\mathbf{0 . 0 1 0}$ | $\mathbf{0 . 0 1 0}$ |
| 1 | 0.069 | 0.141 | 0.141 | 0.122 | 0.017 | 0.015 | 0.000 | 0.045 | 0.015 | $\mathbf{0 . 0 6 6}$ | $\mathbf{0 . 0 4 5}$ |
| 2 | 0.098 | 0.165 | 0.165 | 0.141 | 0.130 | 0.129 | 0.125 | 0.135 | 0.125 | $\mathbf{0 . 1 2 8}$ | $\mathbf{0 . 1 2 8}$ |
| 3 | 0.139 | 0.182 | 0.182 | 0.154 | 0.149 | 0.137 | 0.142 | 0.152 | 0.141 | $\mathbf{0 . 1 4 5}$ | $\mathbf{0 . 1 4 6}$ |
| 4 | 0.155 | 0.203 | 0.203 | 0.166 | 0.172 | 0.160 | 0.175 | 0.175 | 0.174 | $\mathbf{0 . 1 7 4}$ | $\mathbf{0 . 1 7 4}$ |
| 5 | 0.192 | 0.210 | 0.210 | 0.183 | 0.165 | 0.175 | 0.169 | 0.178 | 0.169 | $\mathbf{0 . 1 7 4}$ | $\mathbf{0 . 1 7 4}$ |
| 6 | 0.213 | 0.236 | 0.236 | 0.200 | 0.181 | 0.217 | 0.189 | 0.210 | 0.190 | $\mathbf{0 . 2 0 5}$ | $\mathbf{0 . 2 0 6}$ |
| 7 | 0.232 | 0.237 | 0.237 | 0.227 | 0.197 | 0.214 | 0.216 | 0.229 | 0.216 | $\mathbf{0 . 2 2 6}$ | $\mathbf{0 . 2 2 6}$ |
| 8 | 0.207 | 0.252 | 0.252 | 0.220 | 0.232 | 0.219 | 0.217 | 0.237 | 0.217 | $\mathbf{0 . 2 3 1}$ | $\mathbf{0 . 2 3 2}$ |
| $9+$ | 0.000 | 0.256 | 0.256 | 0.298 | 0.000 | 0.224 | 0.223 | 0.258 | 0.223 | $\mathbf{0 . 2 5 0}$ | $\mathbf{0 . 2 5 0}$ |

Table 2.2.3: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea in 2006. Mean length-at-age (cm) in the catch, by quarter and division.

|  | IIIa <br> NSAS | IVa(E) <br> all | IVa(E) | IVa(W) | IVb |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |$\quad$ IVc $\quad$ VIId |  |
| ---: | :--- |
| IVb |$\quad$|  |
| ---: |
| VIId |

Quarters: 1-4

| 0 | n.d. | n.d. | 0.0 | - |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | n.d. | n.d. | 0.0 | 0.0 |  |
| 2 | n.d. | n.d. | 0.0 | 0.0 |  |
| 3 | n.d. | n.d. | 0.0 | 0.0 |  |
| 4 | n.d. | n.d. | 0.0 | 0.0 |  |
| 5 | n.d. | n.d. | 0.0 | 0.0 |  |
| 6 | n.d. | n.d. | 0.0 | 0.0 |  |
| 7 | n.d. | n.d. | 0.0 | 0.0 |  |
| 8 | n.d. | n.d. | 0.0 | 0.0 |  |
| $9+$ | n.d. |  |  | 0.0 | 0.0 |

Quarter: 1

| 0 | n.d. | n.d. | - | - |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | n.d. | n.d. | 0.0 | 0.0 |
| 2 | n.d. | n.d. | 0.0 | 0.0 |
| 3 | n.d. | n.d. | 0.0 | 0.0 |
| 4 | n.d. | n.d. | 0.0 | 0.0 |
| 5 | n.d. | n.d. | 0.0 | 0.0 |
| 6 | n.d. | n.d. | 0.0 | 0.0 |
| 7 | n.d. | n.d. | 0.0 | 0.0 |
| 8 | n.d. | n.d. | 0.0 | 0.0 |
| $9+$ | n.d. | n.d. | 0.0 | - |
|  |  |  |  |  |

Quarter: 2

| 0 | n.d. | n.d. | - | - |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | n.d. | n.d. | 0.0 | 0.0 |
| 2 | n.d. | n.d. | 0.0 | 0.0 |
| 3 | n.d. | n.d. | 0.0 | 0.0 |
| 4 | n.d. | n.d. | 0.0 | 0.0 |
| 5 | n.d. | n.d. | 0.0 | 0.0 |
| 6 | n.d. | n.d. | 0.0 | 0.0 |
| 7 | n.d. | n.d. | 0.0 | 0.0 |
| 8 | n.d. | n.d. | 0.0 | 0.0 |
| $9+$ | n.d. | n.d. | 0.0 | - |

Quarter: 3

| 0 | n.d. | n.d. | 0.0 | - |
| :--- | :--- | :--- | :--- | :--- | ---: |
| 1 | n.d. | n.d. | 0.0 | \#DIV/0! |
| 2 | n.d. | n.d. | 0.0 | 0.0 |
| 3 | n.d. | n.d. | 0.0 | 0.0 |
| 4 | n.d. | n.d. | 0.0 | 0.0 |
| 5 | n.d. | n.d. | 0.0 | 0.0 |
| 6 | n.d. | n.d. | 0.0 | 0.0 |
| 7 | n.d. | n.d. | 0.0 | 0.0 |
| 8 | n.d. | n.d. | 0.0 | 0.0 |
| $9+$ | n.d. | n.d. | 0.0 | 0.0 |

Quarter: 4

| 0 | n.d. | n.d. | 0.0 | - |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | n.d. | n.d. | 0.0 | 0.0 |
| 2 | n.d. | n.d. | 0.0 | 0.0 |
| 3 | n.d. | n.d. | 0.0 | 0.0 |
| 4 | n.d. | n.d. | 0.0 | 0.0 |
| 5 | n.d. | n.d. | 0.0 | 0.0 |
| 6 | n.d. | n.d. | 0.0 | 0.0 |
| 7 | n.d. | n.d. | 0.0 | 0.0 |
| 8 | n.d. | n.d. | 0.0 | 0.0 |
| $9+$ | n.d. | n.d. | 0.0 | 0.0 |

Table 2.2.4: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea in 2006. Catches (tonnes) at-age (SOP figures), by quarter and division.

| WR | $\begin{array}{r} \text { IIIa } \\ \text { NSAS } \end{array}$ | $\begin{array}{r} \hline \mathrm{IVa}(\mathrm{E}) \\ \text { all } \end{array}$ | $\begin{aligned} & \hline \text { IVa(E) } \\ & \text { WBSS } \end{aligned}$ | $\begin{array}{r} \hline \text { IVa(E) } \\ \text { NSAS } \\ \text { only } \end{array}$ | IVa(W) | IVb | IVc | VIId | $\begin{array}{r} \text { IVa \& } \\ \text { IVb } \\ \text { NSAS } \end{array}$ | $\begin{array}{r} \text { IVc \& } \\ \text { VIId } \end{array}$ | Total NSAS | Herring caught in the North Sea |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarters: 1-4 |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0.5 | 0.0 | 0.0 | 0.0 | 0.1 | 8.4 | 0.0 | 0.1 | 8.4 | 0.1 | 9.0 | 8.5 |
| 1 | 8.1 | 0.0 | 0.0 | 0.0 | 0.4 | 0.8 | 0.6 | 1.1 | 1.2 | 1.7 | 10.9 | 2.9 |
| 2 | 4.0 | 5.8 | 0.4 | 5.4 | 16.7 | 10.2 | 1.3 | 9.1 | 32.4 | 10.4 | 46.7 | 43.2 |
| 3 | 1.2 | 9.3 | 1.3 | 8.0 | 19.9 | 8.5 | 0.5 | 6.8 | 36.3 | 7.3 | 44.8 | 45.0 |
| 4 | 0.5 | 20.8 | 2.4 | 18.4 | 40.8 | 14.9 | 0.3 | 5.3 | 74.1 | 5.5 | 80.1 | 82.1 |
| 5 | 0.6 | 34.2 | 4.0 | 30.2 | 94.1 | 33.4 | 0.5 | 22.6 | 157.7 | 23.2 | 181.4 | 184.8 |
| 6 | 0.1 | 11.1 | 1.0 | 10.1 | 34.5 | 7.8 | 0.1 | 2.9 | 52.5 | 3.0 | 55.6 | 56.5 |
| 7 | 0.1 | 14.3 | 1.0 | 13.3 | 34.8 | 7.8 | 0.0 | 3.1 | 55.9 | 3.1 | 59.1 | 60.0 |
| 8 | 0.0 | 4.5 | 0.4 | 4.1 | 7.5 | 2.2 | 0.0 | 1.1 | 13.8 | 1.2 | 14.9 | 15.4 |
| 9+ | 0.0 | 3.3 | 0.2 | 3.1 | 5.9 | 1.4 | 0.0 | 1.2 | 10.3 | 1.2 | 11.6 | 11.7 |
| Sum | 15.0 | 103.2 | 10.7 | 92.5 | 254.6 | 95.4 | 3.3 | 53.4 | 442.6 | 56.7 | 514.3 | 510.0 |

## Quarter: 1

|  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | ---: | ---: | ---: | ---: |
| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.1 | $\mathbf{0 . 1}$ |  |
| 1 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.6 | 1.1 | 0.2 | 1.7 | $\mathbf{0 . 9}$ | $\mathbf{1 . 9}$ |
| 2 | 1.9 | 0.3 | 0.0 | 0.2 | 0.4 | 3.3 | 0.3 | 0.2 | 4.0 | 0.5 | $\mathbf{6 . 3}$ | $\mathbf{4 . 5}$ |
| 3 | 0.7 | 0.7 | 0.1 | 0.6 | 2.5 | 0.5 | 0.1 | 1.7 | 3.6 | 1.8 | $\mathbf{6 . 0}$ | $\mathbf{5 . 5}$ |
| 4 | 0.2 | 1.8 | 0.0 | 1.7 | 5.3 | 0.1 | 0.0 | 1.2 | 7.1 | 1.3 | $\mathbf{8 . 6}$ | $\mathbf{8 . 4}$ |
| 5 | 0.4 | 2.8 | 0.0 | 2.8 | 7.8 | 0.2 | 0.2 | 7.2 | 10.8 | 7.4 | $\mathbf{1 8 . 6}$ | $\mathbf{1 8 . 2}$ |
| 6 | 0.1 | 0.2 | 0.0 | 0.2 | 2.2 | 0.0 | 0.0 | 0.8 | 2.4 | 0.9 | $\mathbf{3 . 3}$ | $\mathbf{3 . 3}$ |
| 7 | 0.0 | 0.6 | 0.0 | 0.6 | 1.9 | 0.0 | 0.0 | 0.5 | 2.4 | 0.5 | $\mathbf{3 . 0}$ | $\mathbf{3 . 0}$ |
| 8 | 0.0 | 0.4 | 0.0 | 0.4 | 0.3 | 0.0 | 0.0 | 0.3 | 0.7 | 0.4 | $\mathbf{1 . 1}$ | $\mathbf{1 . 1}$ |
| $9+$ | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.6 | 0.2 | 0.6 | $\mathbf{0 . 8}$ | $\mathbf{0 . 8}$ |
| Sum | $\mathbf{4 . 2}$ | $\mathbf{6 . 8}$ | $\mathbf{0 . 2}$ | $\mathbf{6 . 6}$ | $\mathbf{2 0 . 5}$ | $\mathbf{4 . 4}$ | $\mathbf{1 . 3}$ | $\mathbf{1 3 . 9}$ | $\mathbf{3 1 . 5}$ | $\mathbf{1 5 . 1}$ | $\mathbf{5 0 . 8}$ | $\mathbf{4 6 . 8}$ |

Quarter: 2

| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | $\mathbf{0 . 0}$ | $\mathbf{0 . 0}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.2 | 0.0 | $\mathbf{0 . 6}$ | $\mathbf{0 . 2}$ |
| 2 | 0.9 | 4.9 | 0.4 | 4.5 | 3.4 | 0.8 | 0.0 | 0.0 | 8.7 | 0.0 | $\mathbf{9 . 6}$ | $\mathbf{9 . 1}$ |
| 3 | 0.0 | 6.6 | 0.7 | 5.9 | 4.7 | 1.5 | 0.0 | 0.0 | 12.1 | 0.0 | $\mathbf{1 2 . 1}$ | $\mathbf{2 5 . 7}$ |
| 4 | 0.0 | 15.7 | 1.9 | 13.8 | 8.6 | 1.5 | 0.0 | 0.0 | 23.8 | 0.0 | $\mathbf{2 3 . 8}$ | $\mathbf{4 9 . 0}$ |
| 5 | 0.0 | 22.9 | 2.8 | 20.1 | 22.6 | 3.6 | 0.0 | 0.0 | 46.2 | 0.0 | $\mathbf{4 6 . 2}$ | $\mathbf{9 . 9}$ |
| 6 | 0.0 | 4.7 | 0.6 | 4.1 | 4.6 | 0.6 | 0.0 | 0.0 | 9.3 | 0.0 | $\mathbf{9 . 3}$ | $\mathbf{1 0 . 2}$ |
| 7 | 0.0 | 4.0 | 0.5 | 3.5 | 5.4 | 0.8 | 0.0 | 0.0 | 9.7 | 0.0 | $\mathbf{9 . 7}$ | $\mathbf{4 . 0}$ |
| 8 | 0.0 | 2.2 | 0.3 | 2.0 | 1.3 | 0.5 | 0.0 | 0.0 | 3.8 | 0.0 | $\mathbf{2 . 0}$ | $\mathbf{1 . 0}$ |
| $9+$ | 0.0 | 0.7 | 0.1 | 0.6 | 1.1 | 0.2 | 0.0 | 0.0 | 1.9 | 0.0 | $\mathbf{1 . 9}$ | $\mathbf{1 2 3 . 0}$ |
| Sum | $\mathbf{1 . 3}$ | $\mathbf{6 1 . 8}$ | $\mathbf{7 . 2}$ | $\mathbf{5 4 . 5}$ | $\mathbf{5 1 . 6}$ | $\mathbf{9 . 5}$ | $\mathbf{0 . 0}$ | $\mathbf{0 . 0}$ | $\mathbf{1 1 5 . 7}$ | $\mathbf{0 . 0}$ | $\mathbf{1 1 7 . 1}$ |  |

## Quarter: 3

| Quarter: $\mathbf{Q}$ |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 0.4 | 0.0 | 0.0 | 0.0 | 0.1 | 6.5 | 0.0 | 0.0 | 6.6 | 0.0 | $\mathbf{6 . 9}$ | $\mathbf{6 . 6}$ |
| 1 | 3.2 | 0.0 | 0.0 | 0.0 | 0.2 | 0.3 | 0.0 | 0.0 | 0.5 | 0.0 | $\mathbf{3 . 6}$ | $\mathbf{1 5 . 0}$ |
| 2 | 1.0 | 0.7 | 0.0 | 0.7 | 11.0 | 3.3 | 0.0 | 0.0 | 15.0 | 0.0 | $\mathbf{1 6 . 1}$ | $\mathbf{1 6 . 8}$ |
| 3 | 0.3 | 1.5 | 0.4 | 1.1 | 11.4 | 3.9 | 0.0 | 0.0 | 16.4 | 0.0 | $\mathbf{1 6 . 7}$ | $\mathbf{3 4 . 7}$ |
| 4 | 0.1 | 1.1 | 0.3 | 0.8 | 22.4 | 11.1 | 0.0 | 0.0 | 34.4 | 0.0 | $\mathbf{3 4 . 5}$ | $\mathbf{7 8 . 3}$ |
| 5 | 0.1 | 3.7 | 0.9 | 2.9 | 56.4 | 18.2 | 0.0 | 0.0 | 77.4 | 0.0 | $\mathbf{7 7 . 5}$ | $\mathbf{3 1 . 1}$ |
| 6 | 0.0 | 1.1 | 0.2 | 0.8 | 25.6 | 4.5 | 0.0 | 0.0 | 30.9 | 0.0 | $\mathbf{3 0 . 9}$ | $\mathbf{3 2 . 2}$ |
| 7 | 0.0 | 1.0 | 0.2 | 0.8 | 25.9 | 5.3 | 0.0 | 0.0 | 32.0 | 0.0 | $\mathbf{3 2 . 0}$ | $\mathbf{6 . 7}$ |
| 8 | 0.0 | 0.6 | 0.1 | 0.5 | 5.2 | 0.9 | 0.0 | 0.0 | 6.6 | 0.0 | $\mathbf{6 . 6}$ | $\mathbf{5 . 8}$ |
| $9+$ | 0.0 | 0.1 | 0.0 | 0.1 | 4.4 | 1.2 | 0.0 | 0.0 | 5.7 | 0.0 | $\mathbf{5 . 7}$ | $\mathbf{2 2 7 . 7}$ |
| Sum | $\mathbf{5 . 2}$ | $\mathbf{9 . 9}$ | $\mathbf{2 . 2}$ | $\mathbf{7 . 7}$ | $\mathbf{1 6 2 . 5}$ | $\mathbf{5 5 . 2}$ | $\mathbf{0 . 0}$ | $\mathbf{0 . 0}$ | $\mathbf{2 2 5 . 4}$ | $\mathbf{0 . 0}$ | $\mathbf{2 3 0 . 6}$ |  |

Quarter: 4

| 0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 1.9 | 0.0 | 0.0 | 1.9 | 0.0 | $\mathbf{2 . 0}$ | $\mathbf{1 . 9}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 3.4 | 0.0 | 0.0 | 0.0 | 0.2 | 0.1 | 0.0 | 0.0 | 0.3 | 0.0 | $\mathbf{0 . 3}$ |  |
| 2 | 0.2 | 0.0 | 0.0 | 0.0 | 2.0 | 2.7 | 1.0 | 8.9 | 4.8 | 9.9 | $\mathbf{1 4 . 8}$ | $\mathbf{1 4 . 6}$ |
| 3 | 0.2 | 0.2 | 0.0 | 0.2 | 1.2 | 2.6 | 0.5 | 5.0 | 4.1 | 5.5 | $\mathbf{9 . 8}$ | $\mathbf{9 . 6}$ |
| 4 | 0.1 | 2.0 | 0.1 | 1.9 | 4.5 | 2.3 | 0.2 | 4.0 | 8.7 | 4.3 | $\mathbf{1 3 . 1}$ | $\mathbf{1 3 . 0}$ |
| 5 | 0.1 | 4.4 | 0.2 | 4.2 | 7.4 | 11.4 | 0.3 | 15.4 | 23.0 | 15.7 | $\mathbf{3 8 . 8}$ | $\mathbf{3 8 . 9}$ |
| 6 | 0.0 | 5.2 | 0.3 | 4.9 | 2.2 | 2.7 | 0.1 | 2.1 | 9.9 | 2.1 | $\mathbf{1 2 . 0}$ | $\mathbf{1 2 . 3}$ |
| 7 | 0.0 | 8.7 | 0.5 | 8.3 | 1.7 | 1.7 | 0.0 | 2.6 | 11.7 | 2.6 | $\mathbf{1 4 . 3}$ | $\mathbf{1 4 . 7}$ |
| 8 | 0.0 | 1.2 | 0.1 | 1.1 | 0.6 | 0.9 | 0.0 | 0.8 | 2.6 | 0.8 | $\mathbf{3 . 4}$ | $\mathbf{3 . 5}$ |
| $9+$ | 0.0 | 2.4 | 0.1 | 2.2 | 0.2 | 0.0 | 0.0 | 0.6 | 2.4 | 0.6 | $\mathbf{3 . 0}$ | $\mathbf{3 . 2}$ |
| Sum | $\mathbf{4 . 4}$ | $\mathbf{2 4 . 2}$ | $\mathbf{1 . 3}$ | $\mathbf{2 2 . 9}$ | $\mathbf{1 9 . 9}$ | $\mathbf{2 6 . 3}$ | $\mathbf{2 . 0}$ | $\mathbf{3 9 . 5}$ | $\mathbf{6 9 . 1}$ | $\mathbf{4 1 . 5}$ | $\mathbf{1 1 5 . 0}$ | $\mathbf{1 1 1 . 9}$ |

Table 2.2.5: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea in 2006. Percentage age composition (based on numbers, 3+ group summarised), by quarter and division.

| WR | $\begin{array}{r} \text { IIIIa } \\ \text { NSAS } \end{array}$ | $\begin{gathered} \overline{\mathrm{IVa}(\mathrm{E})} \\ \text { all } \end{gathered}$ | IVa(E) WBSS | IVa(E) NSAS only | IVa(W) | IVb | IVc | VIld | $\begin{gathered} \hline \text { TVa\& } \\ \text { IVb } \end{gathered}$ | IVc \& VIId | Total NSAS | Herring caught in the North Sea |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| WR |  |  |  |  |  |  |  |  |  |  |  |  |

Quarters: 1-4

| 0 | 13.9\% | 0.0\% | 0.0\% | 0.0\% | 0.4\% | 59.8\% | 0.6\% | 0.2\% | 26.1\% | 0.3\% | 22.6\% | 22.9\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 59.3\% | 0.1\% | 0.2\% | 0.1\% | 0.3\% | 2.7\% | 46.6\% | 2.4\% | 1.3\% | 7.3\% | 5.7\% | 2.0\% |
| 2 | 19.8\% | 8.1\% | 5.7\% | 8.4\% | 8.9\% | 7.5\% | 30.9\% | 20.2\% | 8.2\% | 21.4\% | 10.3\% | 9.6\% |
| 3 | 4.0\% | 10.8\% | 14.1\% | 10.4\% | 9.8\% | 4.3\% | 9.0\% | 15.0\% | 7.5\% | 14.3\% | 8.0\% | 8.4\% |
| 4 | 1.3\% | 21.9\% | 22.5\% | 21.8\% | 17.3\% | 6.2\% | 3.8\% | 9.5\% | 13.2\% | 8.8\% | 12.0\% | 12.9\% |
| 5 | 1.3\% | 33.7\% | 36.0\% | 33.4\% | 37.3\% | 13.1\% | 7.5\% | 41.1\% | 26.1\% | 37.3\% | 25.7\% | 27.6\% |
| 6 | 0.2\% | 9.0\% | 8.2\% | 9.1\% | 11.5\% | 2.8\% | 1.0\% | 4.5\% | 7.3\% | 4.1\% | 6.5\% | 7.0\% |
| 7 | 0.1\% | 11.0\% | 8.5\% | 11.3\% | 10.6\% | 2.5\% | 0.2\% | 4.2\% | 7.2\% | 3.7\% | 6.4\% | 6.8\% |
| 8 | 0.1\% | 3.4\% | 3.4\% | 3.4\% | 2.3\% | 0.7\% | 0.1\% | 1.5\% | 1.8\% | 1.3\% | 1.6\% | 1.8\% |
| 9+ | 0.0\% | 2.2\% | 1.5\% | 2.3\% | 1.6\% | 0.4\% | 0.2\% | 1.5\% | 1.2\% | 1.4\% | 1.1\% | 1.2\% |
| Sum 3+ | 7.1\% | 91.9\% | 94.1\% | 91.6\% | 90.4\% | 30.0\% | 21.8\% | 77.2\% | 64.4\% | 71.0\% | 61.3\% | 65.6\% |

Quarter: 1

| 0 | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.4\% | 0.1\% | 0.8\% | 0.1\% | 0.6\% | 0.3\% | 0.3\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 57.2\% | 0.8\% | 6.3\% | 0.5\% | 0.3\% | 15.0\% | 68.6\% | 7.6\% | 4.0\% | 20.7\% | 18.5\% | 10.0\% |
| 2 | 30.7\% | 5.7\% | 22.1\% | 4.9\% | 2.9\% | 74.2\% | 21.7\% | 1.6\% | 21.3\% | 5.9\% | 18.4\% | 15.7\% |
| 3 | 7.1\% | 11.7\% | 48.9\% | 10.0\% | 13.8\% | 6.9\% | 2.1\% | 16.8\% | 11.4\% | 13.7\% | 11.3\% | 12.4\% |
| 4 | 1.8\% | 27.1\% | 7.5\% | 28.0\% | 27.8\% | 0.7\% | 1.0\% | 10.0\% | 21.0\% | 8.1\% | 13.7\% | 16.3\% |
| 5 | 2.7\% | 39.4\% | 11.0\% | 40.7\% | 36.9\% | 2.3\% | 5.3\% | 51.6\% | 28.8\% | 41.7\% | 28.0\% | 33.4\% |
| 6 | 0.3\% | 2.5\% | 0.7\% | 2.6\% | 9.2\% | 0.2\% | 0.5\% | 4.8\% | 5.7\% | 3.9\% | 4.2\% | 5.0\% |
| 7 | 0.1\% | 7.5\% | 2.1\% | 7.8\% | 7.5\% | 0.2\% | 0.3\% | 2.8\% | 5.7\% | 2.3\% | 3.7\% | 4.4\% |
| 8 | 0.1\% | 4.9\% | 1.4\% | 5.1\% | 1.1\% | 0.1\% | 0.2\% | 1.6\% | 1.5\% | 1.3\% | 1.2\% | 1.4\% |
| 9+ | 0.0\% | 0.3\% | 0.1\% | 0.3\% | 0.6\% | 0.0\% | 0.2\% | 2.4\% | 0.4\% | 1.9\% | 0.8\% | 0.9\% |
| Sum 3+ | 12.0\% | 93.5\% | 71.6\% | 94.5\% | 96.8\% | 10.4\% | 9.7\% | 90.0\% | 74.6\% | 72.8\% | 62.9\% | 73.9\% |

Quarter: 2

| 0 | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.1\% | 0.8\% | 0.0\% | 0.1\% | 0.0\% | 0.0\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 46.7\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 8.3\% | 58.2\% | 7.6\% | 0.9\% | 55.7\% | 2.3\% | 0.8\% |
| 2 | 53.0\% | 10.3\% | 6.8\% | 10.8\% | 9.1\% | 11.4\% | 14.7\% | 1.6\% | 10.1\% | 14.1\% | 11.4\% | 9.9\% |
| 3 | 0.2\% | 12.4\% | 12.0\% | 12.5\% | 11.5\% | 17.7\% | 11.0\% | 16.8\% | 12.6\% | 11.2\% | 12.3\% | 12.6\% |
| 4 | 0.1\% | 26.1\% | 27.4\% | 25.9\% | 17.0\% | 14.3\% | 2.8\% | 10.0\% | 21.1\% | 3.1\% | 20.4\% | 21.4\% |
| 5 | 0.0\% | 36.1\% | 37.9\% | 35.8\% | 42.8\% | 33.2\% | 11.3\% | 51.6\% | 38.4\% | 13.3\% | 37.2\% | 38.3\% |
| 6 | 0.0\% | 6.5\% | 6.8\% | 6.4\% | 7.9\% | 4.8\% | 0.9\% | 4.8\% | 6.8\% | 1.1\% | 6.6\% | 6.8\% |
| 7 | 0.0\% | 5.1\% | 5.4\% | 5.1\% | 8.0\% | 6.0\% | 0.6\% | 2.8\% | 6.4\% | 0.7\% | 6.2\% | 6.3\% |
| 8 | 0.0\% | 2.7\% | 2.9\% | 2.7\% | 2.1\% | 3.3\% | 0.2\% | 1.6\% | 2.5\% | 0.3\% | 2.5\% | 2.6\% |
| 9+ | 0.0\% | 0.8\% | 0.8\% | 0.8\% | 1.6\% | 1.0\% | 0.3\% | 2.4\% | 1.1\% | 0.4\% | 1.1\% | 1.1\% |
| Sum 3+ | 0.3\% | 89.7\% | 93.2\% | 89.2\% | 90.9\% | 80.3\% | 27.0\% | 90.0\% | 89.0\% | 30.1\% | 86.3\% | 89.2\% |

Quarter: 3

| 0 | 32.9\% | 0.0\% | 0.0\% | 0.0\% | 0.7\% | 70.4\% | 0.0\% | 0.0\% | 37.8\% | 0.0\% | 37.6\% | 37.6\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 49.8\% | 0.0\% | 0.0\% | 0.0\% | 0.2\% | 1.9\% | 0.0\% | 0.0\% | 1.1\% | 0.0\% | 3.3\% | 1.1\% |
| 2 | 12.4\% | 9.9\% | 0.5\% | 12.5\% | 9.4\% | 3.1\% | 0.0\% | 28.0\% | 6.1\% | 28.0\% | 6.4\% | 6.0\% |
| 3 | 3.1\% | 17.5\% | 22.1\% | 16.2\% | 8.8\% | 2.8\% | 0.0\% | 14.5\% | 5.8\% | 14.5\% | 5.6\% | 5.9\% |
| 4 | 1.0\% | 11.5\% | 12.3\% | 11.3\% | 14.5\% | 6.8\% | 0.0\% | 8.7\% | 10.3\% | 8.7\% | 9.9\% | 10.3\% |
| 5 | 0.4\% | 38.8\% | 41.4\% | 38.1\% | 35.7\% | 9.6\% | 0.0\% | 36.3\% | 21.8\% | 36.3\% | 20.8\% | 21.9\% |
| 6 | 0.2\% | 8.2\% | 8.7\% | 8.0\% | 13.5\% | 2.2\% | 0.0\% | 4.5\% | 7.3\% | 4.5\% | 7.0\% | 7.4\% |
| 7 | 0.1\% | 8.2\% | 8.7\% | 8.0\% | 12.8\% | 2.4\% | 0.0\% | 5.0\% | 7.2\% | 5.0\% | 6.8\% | 7.2\% |
| 8 | 0.0\% | 5.0\% | 5.3\% | 4.9\% | 2.6\% | 0.4\% | 0.0\% | 1.8\% | 1.5\% | 1.8\% | 1.4\% | 1.5\% |
| 9+ | 0.0\% | 0.9\% | 1.0\% | 0.9\% | 1.9\% | 0.5\% | 0.0\% | 1.3\% | 1.2\% | 1.3\% | 1.1\% | 1.2\% |
| Sum 3+ | 5.0\% | 90.1\% | 99.5\% | 87.5\% | 89.7\% | 24.7\% | 0.0\% | 72.0\% | 55.0\% | 72.0\% | 52.7\% | 55.3\% |

Quarter: 4

| 0 | 12.8\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 54.6\% | 1.8\% | 0.0\% | 33.6\% | 0.1\% | 22.0\% | 22.7\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 78.5\% | 0.0\% | 0.0\% | 0.0\% | 1.6\% | 1.5\% | 0.0\% | 0.0\% | 1.3\% | 0.0\% | 6.5\% | 0.8\% |
| 2 | 3.4\% | 0.2\% | 0.7\% | 0.2\% | 12.4\% | 6.2\% | 50.7\% | 28.5\% | 6.4\% | 29.7\% | 13.2\% | 14.0\% |
| 3 | 2.4\% | 1.2\% | 2.9\% | 1.1\% | 7.0\% | 5.2\% | 23.4\% | 14.2\% | 4.9\% | 14.7\% | 7.6\% | 8.0\% |
| 4 | 1.3\% | 9.4\% | 9.2\% | 9.4\% | 24.0\% | 3.9\% | 9.8\% | 9.2\% | 9.0\% | 9.2\% | 8.5\% | 9.1\% |
| 5 | 1.1\% | 20.0\% | 19.7\% | 20.0\% | 35.6\% | 20.4\% | 12.1\% | 36.4\% | 23.5\% | 35.0\% | 25.3\% | 27.2\% |
| 6 | 0.1\% | 21.1\% | 20.6\% | 21.1\% | 9.6\% | 4.5\% | 2.0\% | 4.4\% | 8.5\% | 4.3\% | 6.6\% | 7.2\% |
| 7 | 0.3\% | 34.9\% | 34.1\% | 34.9\% | 6.5\% | 2.6\% | 0.1\% | 4.8\% | 9.2\% | 4.5\% | 7.2\% | 7.7\% |
| 8 | 0.1\% | 4.4\% | 4.3\% | 4.4\% | 2.6\% | 1.1\% | 0.0\% | 1.5\% | 2.0\% | 1.4\% | 1.7\% | 1.8\% |
| 9+ | 0.0\% | 8.8\% | 8.6\% | 8.8\% | 0.5\% | 0.0\% | 0.0\% | 1.1\% | 1.7\% | 1.1\% | 1.4\% | 1.5\% |
| Sum 3+ | 5.4\% | 99.7\% | 99.3\% | 99.8\% | 85.9\% | 37.7\% | 47.4\% | 71.5\% | 58.8\% | 70.2\% | 58.3\% | 62.5\% |

Table 2.2.6: Total catch of herring caught in the North Sea and Div. IIIa: North Sea autumn spawners (NSAS). Catch in numbers (millions) at mean weight-at-age (kg) by fleet, and SOP catches ('000 t). SOP catch might deviate from reported catch as used for the assessment.

| 2003 | Fleet A |  | Fleet B |  | Fleet C |  | Fleet D |  | TOTAL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total Winter rings | Mean |  | Mean |  | Mean |  | Mean |  | Mean |  |
|  | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight |
| 0 | 1.7 | 0.038 | 345.8 | 0.013 | 1.9 | 0.013 | 19.7 | 0.021 | 369.1 | 0.014 |
| 1 | 59.2 | 0.078 | 112.8 | 0.030 | 167.5 | 0.054 | 277.5 | 0.021 | 617.0 | 0.037 |
| 2 | 952.9 | 0.115 | 69.2 | 0.048 | 142.1 | 0.073 | 40.2 | 0.048 | 1,204.5 | 0.104 |
| 3 | 502.0 | 0.158 | 1.9 | 0.123 | 12.4 | 0.124 | 0.7 | 0.099 | 516.9 | 0.157 |
| 4 | 799.1 | 0.174 | 4.4 | 0.133 | 16.0 | 0.151 | 0.2 | 0.128 | 819.7 | 0.173 |
| 5 | 240.5 | 0.185 | 0.4 | 0.162 | 1.8 | 0.163 | 0.0 | 0.174 | 242.7 | 0.184 |
| 6 | 104.7 | 0.204 | 0.4 | 0.173 | 1.1 | 0.193 | 0.1 | 0.152 | 106.2 | 0.204 |
| 7 | 118.8 | 0.221 | 0.5 | 0.178 | 1.2 | 0.214 | 0.0 | 0.244 | 120.5 | 0.221 |
| 8 | 36.8 | 0.232 | 0.1 | 0.178 | 0.2 | 0.187 | 0.0 | 0.180 | 37.1 | 0.232 |
| 9+ | 8.3 | 0.253 |  |  |  |  |  |  | 8.3 | 0.253 |
| TOTAL | 2,824.0 |  | 535.5 |  | 344.1 |  | 338.4 |  | 4,041.9 |  |
| SOP catch | 434.8 |  | 12.3 |  | 24.1 |  | 8.4 |  | 479.6 |  |
|  |  |  |  |  | Figures for A fleet include 3809 t unsampled bycatch in the industrial fishery |  |  |  |  |  |
| 2004 | Fleet A |  | Fleet B |  | Fleet C |  | Fleet D |  | TOTAL |  |
| Total | Numbers $\begin{gathered}\text { Mean } \\ \text { Weight }\end{gathered}$ |  | $\begin{array}{cc}\text { Mean } \\ \text { Numbers } & \text { Weight }\end{array}$ |  | Numbers $\begin{gathered}\text { Mean } \\ \text { Weight }\end{gathered}$ |  | $\begin{array}{cc}\text { Mean } \\ \text { Numbers } & \text { Weight }\end{array}$ |  | Numbers $\begin{gathered}\text { Mean } \\ \text { Weight }\end{gathered}$ |  |
| Winter rings |  |  |  |  |  |  |  |  |  |  |
| 0 |  |  | 627.2 | 0.013 | 13.2 | 0.024 | 75.2 | 0.022 | 715.6 | 0.014 |
| 1 | 2.7 | 0.073 | 133.0 | 0.025 | 18.8 | 0.060 | 52.1 | 0.054 | 206.7 | 0.036 |
| 2 | 252.9 | 0.121 | 5.9 | 0.039 | 114.2 | 0.069 | 65.7 | 0.073 | 438.8 | 0.099 |
| 3 | 1298.6 | 0.138 | 6.8 | 0.096 | 12.0 | 0.120 | 8.7 | 0.121 | 1,326.1 | 0.137 |
| 4 | 510.6 | 0.183 | 2.9 | 0.137 | 4.4 | 0.138 | 1.6 | 0.147 | 519.5 | 0.182 |
| 5 | 714.6 | 0.206 | 1.9 | 0.175 | 8.7 | 0.149 | 1.0 | 0.171 | 726.2 | 0.205 |
| 6 | 168.6 | 0.221 | 0.8 | 0.168 | 1.6 | 0.169 | 0.2 | 0.185 | 171.1 | 0.220 |
| 7 | 99.1 | 0.229 | 0.2 | 0.217 | 1.9 | 0.187 | 0.1 | 0.183 | 101.2 | 0.228 |
| 8 | 69.7 | 0.241 | 0.5 | 0.232 | 0.8 | 0.178 | 0.0 | 0.213 | 71.1 | 0.241 |
| 9+ | 22.0 | 0.265 |  |  |  |  |  |  | 22.0 | 0.265 |
| TOTAL | 3,139.0 |  | 779.1 |  | 175.7 |  | 204.7 |  | 4,298.4 |  |
| SOP catch | 532.8 |  | 13.6 |  | 13.4 |  | 10.8 |  | 570.6 |  |
|  |  |  |  |  | Figures for A fleet include 4984 t unsampled bycatch in the industrial fishery |  |  |  |  |  |
| 2005 | Fleet A |  | Fleet B |  | Fleet C |  | Fleet D |  | TOTAL |  |
| Total | Mean |  | Mean |  | Mean |  | Mean |  | Mean |  |
| Winter rings | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight |
| 0 | 0.4 | 0.119 | 918.7 | 0.011 | 11.3 | 0.027 | 85.1 | 0.015 | 1,015.6 | 0.011 |
| 1 | 42.3 | 0.088 | 365.8 | 0.033 | 174.6 | 0.065 | 132.9 | 0.032 | 715.5 | 0.044 |
| 2 | 196.3 | 0.122 | 0.0 | 0.000 | 115.9 | 0.072 | 43.3 | 0.068 | 355.4 | 0.099 |
| 3 | 469.5 | 0.155 | 0.0 | 0.000 | 12.4 | 0.106 | 3.7 | 0.105 | 485.7 | 0.153 |
| 4 | 1313.0 | 0.166 | 0.0 | 0.000 | 4.7 | 0.154 | 0.6 | 0.158 | 1,318.4 | 0.166 |
| 5 | 477.6 | 0.208 | 0.0 | 0.000 | 2.1 | 0.175 | 0.2 | 0.157 | 479.9 | 0.208 |
| 6 | 573.6 | 0.223 | 0.0 | 0.000 | 1.9 | 0.189 | 0.3 | 0.160 | 575.9 | 0.223 |
| 7 | 114.7 | 0.240 | 0.0 | 0.000 | 0.3 | 0.216 | 0.2 | 0.178 | 115.2 | 0.240 |
| 8 | 107.8 | 0.266 | 0.0 | 0.000 | 0.2 | 0.209 | 0.0 | 0.000 | 108.0 | 0.266 |
| 9+ | 39.1 | 0.265 | 0.0 | 0.000 |  |  |  |  | 39.1 | 0.265 |
| TOTAL | 3,334.2 |  | 1,284.5 |  | 323.5 |  | 266.4 |  | 5,208.7 |  |
| SOP catch | 611.7 |  | 21.8 |  | 22.9 |  | 9.0 |  | 665.4 |  |


| 2006 | Fleet A |  | Fleet B |  | Fleet C |  | Fleet D |  | TOTAL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total |  | Mean |  | Mean |  | Mean |  | Mean |  | Mean |
| Winter rings | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight |
| 0 | 7.6 | 0.065 | 835.9 | 0.010 | 6.0 | 0.020 | 29.1 | 0.013 | 878.6 | 0.010 |
| 1 | 14.3 | 0.111 | 57.8 | 0.023 | 93.3 | 0.068 | 56.8 | 0.030 | 222.2 | 0.049 |
| 2 | 334.1 | 0.127 | 20.3 | 0.044 | 42.1 | 0.081 | 8.1 | 0.069 | 404.5 | 0.117 |
| 3 | 308.2 | 0.145 | 1.0 | 0.119 | 7.3 | 0.119 | 2.9 | 0.113 | 319.4 | 0.144 |
| 4 | 471.8 | 0.172 | 3.8 | 0.153 | 2.4 | 0.141 | 0.8 | 0.137 | 478.8 | 0.172 |
| 5 | 1012.6 | 0.181 | 4.7 | 0.160 | 2.1 | 0.184 | 1.2 | 0.188 | 1,020.6 | 0.181 |
| 6 | 257.5 | 0.220 | 0.0 | 0.000 | 0.4 | 0.188 | 0.1 | 0.197 | 258.1 | 0.219 |
| 7 | 253.3 | 0.237 | 0.0 | 0.000 | 0.3 | 0.213 | 0.1 | 0.225 | 253.7 | 0.237 |
| 8 | 64.6 | 0.235 | 0.5 | 0.214 | 0.1 | 0.206 | 0.0 | 0.209 | 65.3 | 0.235 |
| 9+ | 44.7 | 0.262 | 0.0 | 0.000 |  |  |  |  | 44.7 | 0.262 |
| TOTAL | 2,768.8 |  | 924.0 |  | 154.1 |  | 99.2 |  | 3,946.0 |  |
| SOP catch |  | 497.5 |  | 11.8 |  | 11.6 |  | 3.4 |  | 524.3 |

Table 2.2.7: Catch at age (numbers in millions) of herring caught in the North Sea, 1992-2006.
SG Rednose's revisions for 1995-2001 are included (see Sect. 2.2.3).

| Year/rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 7874 | 705 | 995 | 424 | 344 | 351 | 370 | 149 | 39 | 24 | 11274 |
| 1993 | 7254 | 1385 | 792 | 614 | 315 | 222 | 230 | 191 | 88 | 42 | 11133 |
| 1994 | 3834 | 497 | 1438 | 504 | 355 | 117 | 98 | 78 | 71 | 46 | 7038 |
| 1995 | 6294 | 484 | 1319 | 818 | 244 | 122 | 57 | 43 | 69 | 29 | 9480 |
| 1996 | 1795 | 645 | 488 | 516 | 170 | 57 | 22 | 9 | 17 | 4 | 3723 |
| 1997 | 364 | 174 | 565 | 428 | 285 | 109 | 31 | 12 | 19 | 6 | 1993 |
| 1998 | 208 | 254 | 1084 | 525 | 267 | 179 | 89 | 14 | 17 | 4 | 2642 |
| 1999 | 968 | 73 | 487 | 1034 | 289 | 134 | 70 | 28 | 10 | 2 | 3096 |
| 2000 | 873 | 194 | 516 | 453 | 636 | 212 | 82 | 36 | 15 | 3 | 3019 |
| 2001 | 1025 | 58 | 678 | 473 | 279 | 319 | 92 | 39 | 18 | 2 | 2982 |
| 2002 | 319 | 490 | 513 | 913 | 294 | 136 | 164 | 47 | 34 | 7 | 2917 |
| 2003 | 347 | 172 | 1022 | 507 | 809 | 244 | 106 | 121 | 37 | 8 | 3375 |
| 2004 | 627 | 136 | 274 | 1333 | 517 | 721 | 170 | 100 | 70 | 22 | 3970 |
| 2005 | 919 | 408 | 203 | 487 | 1326 | 480 | 577 | 116 | 108 | 39 | 4664 |
| 2006 | 844 | 72 | 354 | 309 | 475 | 1017 | 257 | 252 | 65 | 44 | 3689 |

Table 2.2.8: Catch at age (numbers in millions) of Baltic Spring spawning Herring taken in the North Sea, and transfered to the assessment of the spring spawning stock in IIIa, 1992-2006.

| Year/rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 |  |  | 0.3 | 9.9 | 11.1 | 8.4 | 8.6 | 2.5 | 0.7 | 0.6 | 42.1 |
| 1993 |  |  | 4.2 | 10.8 | 12.3 | 8.4 | 5.9 | 4.7 | 1.7 | 1.0 | 49.0 |
| 1994 |  |  | 8.8 | 28.2 | 16.3 | 11.0 | 8.6 | 3.4 | 3.2 | 0.7 | 80.2 |
| 1995 |  |  | 22.4 | 11.0 | 14.9 | 4.0 | 2.9 | 1.9 | 0.7 | 0.0 | 57.8 |
| 1996 |  |  | 0.0 | 2.8 | 0.8 | 0.4 | 0.1 | 0.1 | 0.3 | 0.0 | 4.5 |
| 1997 |  |  | 2.2 | 1.3 | 1.5 | 0.4 | 0.2 | 0.1 | 0.2 | 0.0 | 5.9 |
| 1998 |  | 5.1 | 9.5 | 12.0 | 10.1 | 6.0 | 3.0 | 0.4 | 0.9 | 0.0 | 47.0 |
| 1999 |  |  | 3.3 | 14.3 | 5.6 | 3.6 | 1.4 | 0.6 | 0.4 | 0.0 | 29.3 |
| 2000 |  |  | 8.2 | 9.8 | 10.2 | 5.7 | 2.5 | 0.6 | 0.7 | 0.1 | 37.6 |
| 2001 |  |  | 11.3 | 10.2 | 6.1 | 7.2 | 2.7 | 1.6 | 0.4 | 0.0 | 39.9 |
| 2002 |  |  | 7.6 | 14.8 | 10.6 | 3.3 | 2.9 | 1.0 | 0.5 | 0.1 | 40.8 |
| 2003 |  |  | 0.0 | 3.1 | 6.0 | 3.5 | 1.2 | 1.3 | 0.5 | 0.1 | 15.7 |
| 2004 |  |  | 15.1 | 27.9 | 3.5 | 4.1 | 1.0 | 0.5 | 0.1 | 0.0 | 52.3 |
| 2005 |  |  | 6.6 | 17.4 | 12.7 | 2.6 | 3.8 | 1.1 | 0.4 | 0.3 | 44.8 |
| 2006 |  | 0.1 | 3.5 | 8.8 | 14.0 | 22.4 | 5.1 | 5.3 | 2.1 | 1.0 | 62.2 |

Table 2.2.9: Catch at age (numbers in millions) of North Sea Autumn Spawners taken in IIIa, and transfered to the assessment of NSAS, 1992-2006. Figures for 1991-1999 were altered in 2001 and 2002, but for 1991-1995 not used $n$ the assessment. SG Rednose's revisions and the revision of 2002 splitting are included (see Sect. 2.2.3).

| Year/rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 2298 | 1409 | 220 | 22 | 10 | 7 | 3 | 1 | 0 | 3971 |
| 1993 | 2795 | 2033 | 238 | 27 | 8 | 4 | 3 | 2 | 1 | 5109 |
| 1994 | 482 | 1087 | 201 | 27 | 6 | 3 | 2 | 0 | 0 | 1807 |
| 1995 | 1145 | 1181 | 147 | 10 | 3 | 1 | 1 | 0 | 0 | 2487 |
| 1996 | 516 | 961 | 154 | 13 | 3 | 1 | 1 | 0 | 0 | 1649 |
| 1997 | 68 | 305 | 125 | 20 | 1 | 1 | 0 | 0 | 0 | 521 |
| 1998 | 51 | 729 | 145 | 25 | 19 | 3 | 3 | 1 | 0 | 977 |
| 1999 | 598 | 231 | 133 | 39 | 10 | 5 | 1 | 1 | 0 | 1017 |
| 2000 | 232 | 978 | 115 | 20 | 21 | 7 | 3 | 1 | 0 | 1377 |
| 2001 | 808 | 557 | 140 | 15 | 1 | 0 | 0 | 0 | 0 | 1521 |
| 2002 | 411 | 345 | 48 | 5 | 1 | 0 | 0 | 0 | 0 | 811 |
| 2003 | 22 | 445 | 182 | 13 | 16 | 2 | 1 | 1 | 0 | 682 |
| 2004 | 88 | 71 | 180 | 21 | 6 | 10 | 2 | 2 | 1 | 380 |
| 2005 | 96 | 307 | 159 | 16 | 5 | 2 | 2 | 0 | 0 | 590 |
| 2006 | 35 | 150 | 50 | 10 | 3 | 3 | 1 | 0 | 0 | 253 |

Table 2.2.10: Catch at age (numbers in millions) of the total North Sea Autumn Spawning stock 1992-2006. Figures for 1991-1999 were altered in 2001 and 2002, but for 1991-1995 not used in the assessment.
SG Rednose's revisions and the revision of 2002 splitting are included (see Sect. 2.2.3).

| Year/rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 10390 | 2470 | 1342 | 445 | 376 | 368 | 383 | 156 | 40 | 23 | 15994 |
| 1993 | 10280 | 4160 | 1305 | 577 | 295 | 210 | 221 | 184 | 86 | 41 | 17358 |
| 1994 | 4437 | 1890 | 1839 | 449 | 332 | 103 | 88 | 74 | 68 | 45 | 9325 |
| 1995 | 7438 | 1665 | 1444 | 817 | 232 | 119 | 55 | 41 | 69 | 29 | 11909 |
| 1996 | 2311 | 1606 | 642 | 526 | 172 | 58 | 23 | 9 | 17 | 4 | 5368 |
| 1997 | 431 | 480 | 688 | 447 | 285 | 109 | 31 | 12 | 19 | 6 | 2507 |
| 1998 | 260 | 978 | 1220 | 538 | 276 | 176 | 89 | 15 | 17 | 4 | 3572 |
| 1999 | 1566 | 304 | 616 | 1059 | 294 | 136 | 69 | 28 | 10 | 2 | 4084 |
| 2000 | 1105 | 1172 | 623 | 463 | 647 | 213 | 82 | 36 | 15 | 2 | 4358 |
| 2001 | 1833 | 614 | 806 | 477 | 274 | 312 | 89 | 37 | 17 | 2 | 4463 |
| 2002 | 730 | 835 | 553 | 903 | 284 | 133 | 161 | 46 | 33 | 7 | 3687 |
| 2003 | 369 | 617 | 1204 | 517 | 820 | 243 | 106 | 120 | 37 | 8 | 4042 |
| 2004 | 716 | 207 | 439 | 1326 | 520 | 726 | 171 | 101 | 71 | 22 | 4298 |
| 2005 | 1016 | 716 | 355 | 486 | 1318 | 480 | 576 | 115 | 108 | 39 | 5209 |
| 2006 | 879 | 222 | 401 | 311 | 465 | 999 | 253 | 249 | 63 | 44 | 3885 |

Table 2.2.11: Comparison of mean weights (kg) at age (rings) in the catch of adult herring in the North Sea (by Div.) and North Sea autumn spawners caught in Div. IIIa in 1996-2006. SG Rednose's revisions for 1995-2001 are included.

| Age (Rings) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Div. | Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | + |
| Illa | 1996 | 0.078 | 0.110 | 0.160 | 0.182 | 0.215 | 0.215 | 0.244 - |  |  |
|  | 1997 | 0.066 | 0.122 | 0.155 | 0.176 | 0.175 | 0.179 | 0.185 - |  |  |
|  | 1998 | 0.078 | 0.118 | 0.163 | 0.180 | 0.197 | 0.179 | 0.226 - |  |  |
|  | 1999 | 0.084 | 0.113 | 0.141 | 0.161 | 0.181 | 0.206 | 0.199 - |  |  |
|  | 2000 | 0.076 | 0.103 | 0.162 | 0.190 | 0.184 | 0.186 | 0.177 - |  |  |
|  | 2001 | 0.073 | 0.105 | 0.128 | 0.133 | 0.224 | 0.170 | 0.192 - |  |  |
|  | 2002 | 0.104 | 0.126 | 0.144 | 0.164 | 0.180 | 0.180 | 0.218 - |  |  |
|  | 2003 | 0.067 | 0.123 | 0.150 | 0.163 | 0.191 | 0.214 | 0.187 - |  |  |
|  | 2004 | 0.070 | 0.121 | 0.141 | 0.152 | 0.170 | 0.187 | 0.178 - |  |  |
|  | 2005 | 0.071 | 0.106 | 0.155 | 0.173 | 0.185 | 0.200 | 0.209 - |  |  |
|  | 2006 | 0.079 | 0.117 | 0.140 | 0.186 | 0.191 | 0.216 | 0.207 - |  |  |
| IVa(E) | 1996 | 0.131 | 0.141 | 0.168 | 0.196 | 0.217 | 0.218 | 0.242 | 0.300 |  |
|  | 1997 | 0.122 | 0.149 | 0.174 | 0.204 | 0.228 | 0.229 | 0.221 | 0.313 |  |
|  | 1998 | 0.114 | 0.148 | 0.171 | 0.199 | 0.219 | 0.237 | 0.269 | 0.233 |  |
|  | 1999 | 0.125 | 0.143 | 0.162 | 0.191 | 0.207 | 0.226 | 0.232 | 0.272 |  |
|  | 2000 | 0.130 | 0.154 | 0.172 | 0.195 | 0.202 | 0.218 | 0.261 | 0.256 |  |
|  | 2001 | 0.121 | 0.148 | 0.165 | 0.177 | 0.197 | 0.220 | 0.262 | 0.238 |  |
|  | 2002 | 0.130 | 0.154 | 0.167 | 0.189 | 0.198 | 0.212 | 0.229 | 0.238 |  |
|  | 2003 | 0.122 | 0.154 | 0.162 | 0.177 | 0.189 | 0.203 | 0.213 | 0.218 |  |
|  | 2004 | 0.119 | 0.133 | 0.171 | 0.185 | 0.212 | 0.192 | 0.218 | 0.252 |  |
|  | 2005 | 0.117 | 0.146 | 0.153 | 0.202 | 0.209 | 0.233 | 0.262 | 0.265 |  |
|  | 2006 | 0.125 | 0.149 | 0.164 | 0.175 | 0.214 | 0.224 | 0.229 | 0.254 |  |
| IVa(W) | 1996 | 0.131 | 0.167 | 0.215 | 0.218 | 0.237 | 0.275 | 0.301 | 0.278 |  |
|  | 1997 | 0.127 | 0.166 | 0.218 | 0.248 | 0.246 | 0.262 | 0.294 | 0.289 |  |
|  | 1998 | 0.130 | 0.170 | 0.205 | 0.244 | 0.263 | 0.270 | 0.308 | 0.314 |  |
|  | 1999 | 0.129 | 0.162 | 0.192 | 0.227 | 0.250 | 0.261 | 0.272 | 0.309 |  |
|  | 2000 | 0.127 | 0.159 | 0.187 | 0.214 | 0.237 | 0.271 | 0.293 | 0.265 |  |
|  | 2001 | 0.138 | 0.168 | 0.193 | 0.222 | 0.235 | 0.266 | 0.285 | 0.296 |  |
|  | 2002 | 0.144 | 0.161 | 0.191 | 0.211 | 0.230 | 0.242 | 0.261 | 0.263 |  |
|  | 2003 | 0.130 | 0.167 | 0.184 | 0.202 | 0.224 | 0.237 | 0.259 | 0.276 |  |
|  | 2004 | 0.131 | 0.155 | 0.193 | 0.220 | 0.242 | 0.251 | 0.246 | 0.299 |  |
|  | 2005 | 0.122 | 0.158 | 0.174 | 0.213 | 0.229 | 0.245 | 0.275 | 0.267 |  |
|  | 2006 | 0.145 | 0.156 | 0.180 | 0.193 | 0.230 | 0.251 | 0.247 | 0.286 |  |
| IVb | 1996 | 0.111 | 0.184 | 0.209 | 0.230 | 0.249 | 0.297 | 0.282 | 0.287 |  |
|  | 1997 | 0.124 | 0.170 | 0.210 | 0.230 | 0.259 | 0.263 | 0.286 | 0.286 |  |
|  | 1998 | 0.117 | 0.162 | 0.203 | 0.216 | 0.243 | 0.218 | 0.311 | 0.307 |  |
|  | 1999 | 0.118 | 0.148 | 0.154 | 0.207 | 0.226 | 0.209 | 0.287 | 0.345 |  |
|  | 2000 | 0.118 | 0.173 | 0.194 | 0.224 | 0.229 | 0.251 | 0.240 | 0.268 |  |
|  | 2001 | 0.105 | 0.150 | 0.176 | 0.188 | 0.199 | 0.206 | 0.244 | 0.275 |  |
|  | 2002 | 0.086 | 0.149 | 0.161 | 0.206 | 0.214 | 0.189 | 0.270 | 0.241 |  |
|  | 2003 | 0.098 | 0.161 | 0.178 | 0.195 | 0.214 | 0.214 | 0.222 | 0.281 |  |
|  | 2004 | 0.118 | 0.143 | 0.186 | 0.214 | 0.234 | 0.239 | 0.297 | 0.308 |  |
|  | 2005 | 0.132 | 0.172 | 0.187 | 0.217 | 0.220 | 0.245 | 0.253 | 0.252 |  |
|  | 2006 | 0.097 | 0.141 | 0.172 | 0.183 | 0.202 | 0.220 | 0.232 | 0.239 |  |
| IVa \& IVb | 1996 | 0.124 | 0.162 | 0.199 | 0.215 | 0.236 | 0.267 | 0.282 | 0.288 |  |
|  | 1997 | 0.125 | 0.161 | 0.202 | 0.233 | 0.245 | 0.254 | 0.264 | 0.291 |  |
|  | 1998 | 0.123 | 0.162 | 0.194 | 0.224 | 0.243 | 0.253 | 0.293 | 0.283 |  |
|  | 1999 | 0.124 | 0.155 | 0.179 | 0.213 | 0.236 | 0.250 | 0.264 | 0.301 |  |
|  | 2000 | 0.125 | 0.162 | 0.185 | 0.210 | 0.227 | 0.258 | 0.275 | 0.263 |  |
|  | 2001 | 0.129 | 0.156 | 0.180 | 0.202 | 0.217 | 0.242 | 0.275 | 0.285 |  |
|  | 2002 | 0.119 | 0.157 | 0.177 | 0.203 | 0.219 | 0.228 | 0.253 | 0.253 |  |
|  | 2003 | 0.113 | 0.163 | 0.178 | 0.190 | 0.210 | 0.225 | 0.239 | 0.255 |  |
|  | 2004 | 0.122 | 0.147 | 0.187 | 0.210 | 0.227 | 0.233 | 0.247 | 0.266 |  |
|  | 2005 | 0.121 | 0.157 | 0.172 | 0.212 | 0.225 | 0.242 | 0.269 | 0.265 |  |
|  | 2006 | 0.123 | 0.150 | 0.174 | 0.187 | 0.222 | 0.239 | 0.238 | 0.269 |  |
| $\overline{\text { IVc \& VIId }}$ | 1996 | 0.121 | 0.143 | 0.159 | 0.185 | 0.194 | 0.203 | 0.155 - |  |  |
|  | 1997 | 0.101 | 0.133 | 0.156 | 0.168 | 0.166 | 0.190 | 0.163 - |  |  |
|  | 1998 | 0.096 | 0.114 | 0.146 | 0.149 | 0.184 | 0.000 | 0.176 - |  |  |
|  | 1999 | 0.116 | 0.139 | 0.159 | 0.189 | 0.198 | 0.217 - | - |  |  |
|  | 2000 | 0.106 | 0.133 | 0.150 | 0.180 | 0.194 | 0.203 - | - |  |  |
|  | 2001 | 0.113 | 0.138 | 0.171 | 0.167 | 0.171 | 0.168 | 0.180 - |  |  |
|  | 2002 | 0.108 | 0.123 | 0.153 | 0.170 | 0.187 | 0.219 | 0.208 - |  |  |
|  | 2003 | 0.103 | 0.127 | 0.144 | 0.168 | 0.176 | 0.188 | 0.200 | 0.227 |  |
|  | 2004 | 0.099 | 0.113 | 0.135 | 0.162 | 0.184 | 0.191 | 0.186 | 0.224 |  |
|  | 2005 | 0.122 | 0.132 | 0.139 | 0.170 | 0.207 | 0.228 | 0.237 | 0.245 |  |
|  | 2006 | 0.119 | 0.125 | 0.153 | 0.152 | 0.178 | 0.205 | 0.209 | 0.219 |  |
| Total | 1996 | 0.123 | 0.157 | 0.189 | 0.205 | 0.212 | 0.262 | 0.280 | 0.288 |  |
| North Sea | 1997 | 0.118 | 0.149 | 0.195 | 0.227 | 0.227 | 0.235 | 0.245 | 0.291 |  |
| Catch | 1998 | 0.119 | 0.146 | 0.185 | 0.219 | 0.239 | 0.253 | 0.288 | 0.283 |  |
|  | 1999 | 0.123 | 0.152 | 0.172 | 0.208 | 0.233 | 0.246 | 0.264 | 0.301 |  |
|  | 2000 | 0.122 | 0.159 | 0.180 | 0.202 | 0.217 | 0.247 | 0.275 | 0.263 |  |
|  | 2001 | 0.118 | 0.149 | 0.177 | 0.198 | 0.213 | 0.238 | 0.267 | 0.288 |  |
|  | 2002 | 0.118 | 0.153 | 0.170 | 0.199 | 0.214 | 0.228 | 0.250 | 0.252 |  |
|  | 2003 | 0.104 | 0.158 | 0.174 | 0.184 | 0.205 | 0.222 | 0.232 | 0.256 |  |
|  | 2004 | 0.100 | 0.138 | 0.183 | 0.201 | 0.216 | 0.228 | 0.246 | 0.272 |  |
|  | 2005 | 0.099 | 0.153 | 0.166 | 0.208 | 0.223 | 0.240 | 0.257 | 0.278 |  |
|  | 2006 | 0.122 | 0.145 | 0.172 | 0.181 | 0.220 | 0.237 | 0.235 | 0.262 |  |

Table 2.2.12: Sampling of commercial landings of herring in the North Sea (Div. IV and VIId) in 2006 by quarter. Sampled catch means the proportion of the reported catch to which sampling was applied. It is limited by $\mathbf{1 0 0} \%$ but might exceed the official landings due to sampling of discards, unallocated and misreported catches. It is not possible to judge the quality of the sampling by this figure alone. Note that only one nation sampled their by-catches in the industrial fishery (Denmark, fleet B). Metiers are each reported combination of nation/fleet/area/quarter.

| Country (fleet) | Quarter | No of metiers | Metiers sampled | Sampled Catch \% | Official Catch | $\begin{array}{r} \text { No. of } \\ \text { samples } \end{array}$ | $\begin{array}{r} \hline \text { No. fish } \\ \text { aged } \\ \hline \end{array}$ | No. fish measured | >1 sample per 1 kt catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 4 | 1 | 0 | 0\% | 3 | 0 | 0 | 0 | n |
| total |  | 1 | 0 | 0\% | 3 | 0 | 0 | 0 | n |
| Denmark (A) | 1 | 4 | 3 | 99\% | 25257 | 11 | 1661 | 1685 | n |
|  | 2 | 3 | 2 | 96\% | 6419 | 8 | 1099 | 1109 | y |
|  | 3 | 3 | 2 | 100\% | 40478 | 44 | 6195 | 6199 | y |
|  | 4 | 3 | 3 | 100\% | 18267 | 13 | 1833 | 1834 | n |
| total |  | 13 | 10 | 99\% | 90421 | 76 | 10788 | 10827 | n |
| Denmark (B) | 1 | 4 | 2 | 96\% | 1425 | 8 | 210 | 210 | y |
|  | 2 | 2 | 1 | 97\% | 282 | 4 | 9 | 11 | y |
|  | 3 | 2 | 1 | 99\% | 6550 | 9 | 515 | 559 | y |
|  | 4 | 3 | 2 | 100\% | 3644 | 8 | 24 | 29 | y |
| total |  | 11 | 6 | 99\% | 11901 | 29 | 758 | 809 | y |
| England and $V$ | 1 | 3 | 0 | 0\% | 995 | 0 | 0 | 0 | n |
|  | 2 | 4 | 0 | 0\% | 3294 | 0 | 0 | 0 | n |
|  | 3 | 2 | 0 | 0\% | 10237 | 0 | 0 | 0 | n |
|  | 4 | 3 | 0 | 0\% | 7672 | 0 | 0 | 0 | n |
| total |  | 12 | 0 | 0\% | 22198 | 0 | 0 | 0 | n |
| Faroe IsI | 1 | 1 | 0 | 0\% | 140 | 0 | 0 | 0 | n |
|  | 3 | 1 | 0 | 0\% | 60 | 0 | 0 | 0 | n |
|  | 4 | 3 | 0 | 0\% | 1585 | 0 | 0 | 0 | n |
| total |  | 5 | 0 | 0\% | 1785 | 0 | 0 | 0 | n |
| France | 1 | 2 | 0 | 0\% | 4308 | 0 | 0 | 0 | n |
|  | 2 | 3 | 0 | 0\% | 4893 | 0 | 0 | 0 | n |
|  | 3 | 3 | 0 | 0\% | 30964 | 0 | 0 | 0 | n |
|  | 4 | 2 | 0 | 0\% | 9310 | 0 | 0 | 0 | n |
| total |  | 10 | 0 | 0\% | 49476 | 0 | 0 | 0 | n |
| Germany | 1 | 1 | 1 | 100\% | 401 | 9 | 487 | 1282 | y |
|  | 2 | 2 | 0 | 0\% | 4431 | 0 | 0 | 0 | n |
|  | 3 | 3 | 1 | 79\% | 12998 | 20 | 662 | 7049 | y |
|  | 4 | 3 | 2 | 83\% | 22584 | 26 | 580 | 9716 | y |
| total |  | 9 | 4 | 72\% | 40414 | 55 | 1729 | 18047 | y |
| Netherlands | 1 | 4 | 2 | 100\% | 4811 | 14 | 350 | 2542 | y |
|  | 2 | 4 | 2 | 100\% | 14250 | 37 | 925 | 5560 | y |
|  | 3 | 2 | 2 | 100\% | 34004 | 72 | 1800 | 7904 | y |
|  | 4 | 4 | 2 | 86\% | 23250 | 8 | 200 | 1310 | n |
| total |  | 14 | 8 | 100\% | 76315 | 131 | 3275 | 17316 | y |
| Northern Irelar | 1 | 1 | 0 | 0\% | 399 | 0 | 0 | 0 | n |
|  | 3 | 1 | 0 | 0\% | 3127 | 0 | 0 | 0 | n |
|  | 4 | 1 | 0 | 0\% | 5 | 0 | 0 | 0 | n |
| total |  | 3 | 0 | 0\% | 3531 | 0 | 0 | 0 | n |
| Norway | 1 | 3 | 1 | 91\% | 3213 | 1 | 50 | 80 | n |
|  | 2 | 3 | 3 | 100\% | 80865 | 28 | 1829 | 3215 | n |
|  | 3 | 3 | 1 | 61\% | 29563 | 3 | 150 | 295 | n |
|  | 4 | 2 | 1 | 95\% | 21720 | 4 | 150 | 212 | n |
| total |  | 11 | 6 | 90\% | 135361 | 36 | 2179 | 3802 | n |
| Scotland | 1 | 2 | 1 | 25\% | 614 | 1 | 50 | 211 | y |
|  | 2 | 4 | 3 | 100\% | 2794 | 17 | 932 | 4006 | y |
|  | 3 | 2 | 1 | 98\% | 44455 | 59 | 3870 | 10518 | y |
|  | 4 | 2 | 0 | 0\% | 566 | 0 | 0 | 0 | n |
| total |  | 10 | 5 | 99\% | 48429 | 77 | 4852 | 14735 | y |
| Sweden | 2 | 3 | 0 | 0\% | 3120 | 0 | 0 | 0 | n |
|  | 3 | 3 | 0 | 0\% | 5836 | 0 | 0 | 0 | n |
|  | 4 | 2 | 0 | 0\% | 1573 | 0 | 0 | 0 | n |
| total |  | 8 | 0 | 0\% | 10529 | 0 | 0 | 0 | n |
| grand total |  | 107 | 78 | 79\% | 490362 | 404 | 23581 | 65536 | n |
| Period total 1 |  | 25 | 10 | 96\% | 41562 | 44 | 2808 | 6010 | y |
| Period total 2 |  | 28 | 11 | 90\% | 120348 | 94 | 4794 | 13901 | n |
| Period total 3 |  | 25 | 8 | 74\% | 218273 | 207 | 13192 | 32524 | n |
| Period total 4 |  | 29 | 10 | 74\% | 110179 | 59 | 2787 | 13101 | n |
| Total for stock 2006 |  | 107 | 39 | 79\% | 490362 | 404 | 23581 | 65536 | n |
| Human Cons. only |  | 96 | 33 | 79\% | 478461 | 375 | 22823 | 64727 | n |
| Total for stock 2004 |  | 100 | 39 | 94\% | 484159 | 519 | 18643 | 93311 | y |
| Total for stock 2005 |  | 102 | 39 | 95\% | 568312 | 438 | 15499 | 89011 | n |
| Human Cons. only 2005 |  | 95 | 35 | 94\% | 546650 | 394 | 14888 | 87114 | n |

Table 2.2.13: Revision of historic catch numbers and mean weights at age in the catch due to incorrect allocation of fish to the plus group and new information on misreporting

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2006$ | SSESSMEN | CATCH NU GE | MBERS AT |  |  |  |  |  |  |  | Tonnes |
| 2001 | 1832691 | 614321 | 806490 | 477456 | 274048 | 311892 | 89298 | 37485 | 17218 | 2360 | 363343 |
| 2002 | 730279 | 835273 | 553042 | 903157 | 283997 | 133206 | 161196 | 46280 | 33355 | 7186 | 370941 |
| 2003 | 369074 | 616986 | 1204451 | 516945 | 819715 | 242669 | 106172 | 120497 | 37075 | 8313 | 472587 |
| 2004 | 715597 | 206658 | 438762 | 1326124 | 519503 | 726235 | 171149 | 101243 | 71100 | 22045 | 567252 |
| 2005 | 1015554 | 715547 | 355438 | 485676 | 1318373 | 479949 | 575851 | 115164 | 107986 | 39113 | 663813 |
| Revised 2001-2004 data |  |  |  |  |  |  |  |  |  |  |  |
| 2001 | 1832691 | 614469 | 842635 | 485628 | 278884 | 321743 | 90918 | 38252 | 17910 | 2692 | 374065 |
| 2002 | 730279 | 837557 | 579592 | 970577 | 292205 | 140701 | 174570 | 48908 | 34620 | 8702 | 394709 |
| 2003 | 369074 | 617021 | 1221992 | 529386 | 835552 | 244780 | 107751 | 123291 | 37671 | 9044 | 482281 |
| 2004 | 715597 | 206648 | 447918 | 1366155 | 543376 | 753231 | 169324 | 104945 | 65341 | 31801 | 587698 |
| 2005 | 1015554 | 715547 | 355453 | 485746 | 1318647 | 479961 | 576154 | 115212 | 88311 | 58497 | 663813 |
| 2006 Assessment mean weights in CATCH |  |  |  |  |  |  |  |  |  |  |  |
| 2001 | 0.012 | 0.048 | 0.117 | 0.149 | 0.177 | 0.197 | 0.212 | 0.237 | 0.267 | 0.286 |  |
| 2002 | 0.012 | 0.037 | 0.116 | 0.151 | 0.169 | 0.198 | 0.214 | 0.228 | 0.25 | 0.253 |  |
| 2003 | 0.014 | 0.037 | 0.104 | 0.157 | 0.173 | 0.184 | 0.204 | 0.221 | 0.232 | 0.253 |  |
| 2004 | 0.014 | 0.036 | 0.099 | 0.138 | 0.182 | 0.200 | 0.216 | 0.227 | 0.245 | 0.272 |  |
| 2005 | 0.011 | 0.044 | 0.099 | 0.153 | 0.166 | 0.208 | 0.222 | 0.239 | 0.266 | 0.265 |  |
| Revised mean weights at age in THE CATCH |  |  |  |  |  |  |  |  |  |  |  |
| 2001 | 0.012 | 0.048 | 0.118 | 0.149 | 0.177 | 0.198 | 0.213 | 0.238 | 0.267 | 0.288 |  |
| 2002 | 0.012 | 0.037 | 0.118 | 0.153 | 0.170 | 0.199 | 0.214 | 0.228 | 0.250 | 0.252 |  |
| 2003 | 0.014 | 0.037 | 0.104 | 0.158 | 0.174 | 0.184 | 0.205 | 0.222 | 0.232 | 0.256 |  |
| 2004 | 0.014 | 0.036 | 0.100 | 0.138 | 0.183 | 0.201 | 0.216 | 0.228 | 0.246 | 0.272 |  |
| 2005 | 0.011 | 0.044 | 0.099 | 0.153 | 0.166 | 0.208 | 0.223 | 0.24 | 0.257 | 0.278 |  |
| Percentage change in catch at AGE |  |  |  |  |  |  |  |  |  |  |  |
| 2001 | 0.00\% | 0.02\% | 4.48\% | 1.71\% | 1.76\% | 3.16\% | 1.81\% | 2.04\% | 4.02\% | 14.07\% | 2.95\% |
| 2002 | 0.00\% | 0.27\% | 4.80\% | 7.46\% | 2.89\% | 5.63\% | 8.30\% | 5.68\% | 3.79\% | 21.10\% | 6.41\% |
| 2003 | 0.00\% | 0.01\% | 1.46\% | 2.41\% | 1.93\% | 0.87\% | 1.49\% | 2.32\% | 1.61\% | 8.79\% | 2.05\% |
| 2004 | 0.00\% | 0.00\% | 2.09\% | 3.02\% | 4.60\% | 3.72\% | -1.07\% | 3.66\% | -8.10\% | 44.25\% | 3.60\% |
| 2005 | 0.00\% | 0.00\% | 0.00\% | 0.01\% | 0.02\% | 0.00\% | 0.05\% | 0.04\% | 18.22\% | 49.56\% | 0.00\% |
| Percentage change in mean weights at age IN THE CATCH |  |  |  |  |  |  |  |  |  |  |  |
| 2001 | 0.00\% | 0.01\% | 0.96\% | 0.33\% | 0.26\% | 0.42\% | 0.25\% | 0.30\% | 0.04\% | 0.55\% |  |
| 2002 | 0.00\% | 0.39\% | 1.56\% | 1.27\% | 0.47\% | 0.31\% | 0.12\% | 0.16\% | -0.16\% | -0.56\% |  |
| 2003 | 0.00\% | 0.02\% | 0.43\% | 0.33\% | 0.29\% | 0.18\% | 0.32\% | 0.31\% | 0.20\% | 1.17\% |  |
| 2004 | 0.00\% | 0.00\% | 0.76\% | 0.57\% | 0.49\% | 0.37\% | 0.17\% | 0.36\% | 0.27\% | 0.24\% |  |
| 2005 | 0.00\% | 0.00\% | 0.00\% | 0.00\% | 0.00\% | 0.00\% | 0.45\% | 0.42\% | -3.38\% | 4.91\% |  |

Table .2.3.1.1: Vessels, areas and cruise dates during the 2006 herring acoustic surveys.

| VESSEL | PERIOD | AREA | RECTANGLES |
| :---: | :---: | :---: | :---: |
| FV Enterprise (SCO) | 1 July - 21 July | $\begin{aligned} & 56^{\circ}-60^{\circ} 30^{\prime} \mathrm{N}, 3^{\circ}-10^{\circ} \\ & \mathrm{W} \end{aligned}$ | $\begin{aligned} & \text { 41E0-E3, 42E0-E3, 43E0- } \\ & \mathrm{E} 3,44 \mathrm{E} 0-\mathrm{E} 3,45 \mathrm{E} 0-\mathrm{E} 4, \\ & 46 \mathrm{E} 2-\mathrm{E} 5,47 \mathrm{E} 2-\mathrm{E} 6,48 \mathrm{E} 3- \\ & \mathrm{E} 6,49 \mathrm{E} 5 \end{aligned}$ |
| Johan Hjort (NOR) | 19 June - 16 July | $\begin{aligned} & 56^{\circ} 30^{\prime} \mathrm{N}-62^{\circ} \mathrm{N}, 2^{\circ}- \\ & 6^{\circ} \mathrm{E} \end{aligned}$ | 42F2-F5, 43F2-F5, 44F2-F5, 45F2-F5, 46F2- F4, 47F2- <br> F4, 48F2-F4, 49F2-F4, 50F2-F4, 51F2-F4, 52F2-F4, plus overlap area A |
| $\begin{aligned} & \hline \text { Scotia } \\ & \text { (SCO) } \end{aligned}$ | 1 July - 21 July | $\begin{aligned} & 57^{\circ}-62^{\circ} \mathrm{N}, 2 / 4^{\circ} \mathrm{W}- \\ & 2^{\circ} \mathrm{E} \end{aligned}$ | $\begin{aligned} & \text { 43E8-F1, 44E6-F1, 45F0- } \\ & \text { F1, 46E6-F1, 47E6-F1, } \\ & \text { 48E6-F1, 49E6-F1, 50E7- } \\ & \text { F1, 51E8-F1, 52E9-F1 } \end{aligned}$ |
| Tridens (NED) | 26 June - 21 July | $53^{\circ} 30^{\prime}-58^{\circ} 30^{\prime} \mathrm{N}$, Eng/Sco to Den/Ger coasts | 38F2-F7, 40E8-F7, 41E7- <br> F7, 42E7- F1, 45E6-E9 |
| Solea (GER) | 29 June - 18 July | $52^{\circ}-56^{\circ} 30^{\prime}$ N, Eng to Den/Ger coasts | $\begin{aligned} & \text { 33F1-F4, 34F2-F4, 35F2-F4, } \\ & \text { 36F0-F7, 37E9- F8, 38E9- } \\ & \text { F1, 39E8-F7 } \end{aligned}$ |
| Dana (DEN) | 25 June - 6 July | Kattegat north of $56^{\circ}+$ Skagerrak and North Sea north of $56^{\circ} 30^{\prime} \mathrm{N}$, east of $6^{\circ} E$ | 41G1-G2, 42F6-F7, 42G0G3, 43F6-G2, 44F6-G1, 45F6, 45F8-G1, 46F9-G0 |

Table 2.3.1.2: Total numbers (millions of fish) and biomass (thousands of tonnes) of North Sea autumn spawning herring in the area surveyed in the acoustic surveys July 2006, with mean weights and mean lengths by age ring.

| AGE ( RING) | NUMBERS <br> (MILLIONS) | BIOMASs <br> $(‘ \mathbf{0 0 0} \mathbf{~ T}$ | MATURITY | WEIGHT <br> (G) | LENGTH <br> (CM) |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 0 | 4621.8 | 42.1 | 0.00 | 9.1 | 10.65 |
| 1 | 6822.8 | 305.2 | 0.00 | 44.7 | 17.92 |
| 2 | 3772.3 | 477.8 | 0.66 | 126.7 | 24.22 |
| 3 | 1997.2 | 315.2 | 0.88 | 157.9 | 25.96 |
| 4 | 2097.5 | 394.3 | 0.98 | 188.0 | 27.19 |
| 5 | 4175.1 | 784.4 | 1.00 | 187.9 | 27.23 |
| 6 | 618.2 | 139.2 | 1.00 | 225.2 | 28.68 |
| 7 | 562.1 | 136.5 | 1.00 | 242.8 | 29.33 |
| 8 | 70.3 | 20.5 | 1.00 | 243.9 | 29.50 |
| $9+$ | 12994.4 | 18.6 | 1.00 | 265.0 | 30.19 |
| Immature | 11827.3 | 2129.9 |  | 38.8 | 15.92 |
| Mature | 24821.7 | 2633.8 |  | 180.1 | 26.83 |
| Total |  |  |  | 106.1 | 21.12 |

Table 2.3.1.3. Revised numbers (millions) and biomass (thousands of tonnes) breakdown by age (winter rings) and maturity obtained for the 2005 International North Sea Herring Acoustic Survey.

| North SEA | NUMBERS <br> (MILLIONS) | BIOMASs <br> (‘000 T) | MATURITY | WEIGHT <br> (G) | LENGTH <br> (CM) |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 5015.9 | 16.0 | 0.00 | 3.2 | 7.9 |
| 1 | 3114.1 | 134.8 | 0.01 | 43.3 | 17.5 |
| 2 | 2055.1 | 276.0 | 0.76 | 134.3 | 24.4 |
| 3 | 3648.5 | 617.8 | 0.96 | 169.3 | 26.1 |
| 4 | 5789.6 | 1040.2 | 0.96 | 179.7 | 26.5 |
| 5 | 1212.9 | 277.1 | 1.00 | 228.5 | 28.5 |
| 6 | 1174.9 | 290.7 | 1.00 | 247.5 | 29.2 |
| 7 | 139.9 | 35.3 | 1.00 | 252.6 | 29.5 |
| 8 | 126.5 | 34.7 | 1.00 | 274.4 | 30.2 |
| $9+$ | 106.7 | 31.5 | 1.00 | 295.1 | 30.7 |
| Immature | 8994.7 | 243.5 |  |  |  |
| Mature | 9890.7 | 1911.1 |  |  |  |
| Total | 22384.3 | 2754.2 |  |  |  |
| 1+ group | 17368.4 |  |  |  |  |

Table 2.3.1.4. Difference in number at age between original and revised estimates for the 2005 International North Sea Herring Acoustic Survey.

| AGE |  | CHANGE IN <br> NUMBER <br> (MILLIONS |
| ---: | ---: | ---: |
| 0 | $0.00 \%$ | 0.0 |
| 1 | $0.05 \%$ | 1.6 |
| 2 | $8.73 \%$ | 164.9 |
| 3 | $6.17 \%$ | 212.1 |
| 4 | $3.21 \%$ | 180.3 |
| 5 | $0.13 \%$ | 1.6 |
| 6 | $0.23 \%$ | 2.7 |
| 7 | $0.00 \%$ | 0.0 |
| 8 | $0.00 \%$ | 0.0 |
| $9+$ | $0.00 \%$ | 0.0 |
| Immature | $0.83 \%$ | 74.2 |
| Mature | $3.27 \%$ | 312.9 |
| Total | $2.58 \%$ | 563.2 |
| $1+$ group | $3.35 \%$ | 563.2 |

Table 2.3.1.5: Estimates of North Sea autumn spawners (millions) at age from acoustic surveys, 1984-2006. For 1984-1986 the estimates are the sum of those from the Division IVa summer survey, the Division IVb autumn survey, and the Divisions IVc, VIId winter survey. The 1987 to 2006 estimates are from the summer survey in Divisions IVa,b and IIIa excluding estimates of Division IIIa/Baltic spring spawners. For 1999 and 2000 the Kattegat was excluded from the results because it was not surveyed.

| AGE <br> (RINGS) | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 551 | 726 | 1,639 | 13,736 | 6,431 | 6,333 | 6,249 | 3,182 | 6,351 | 10,399 | 3,646 | 4,202 |
| 2 | 3,194 | 2,789 | 3,206 | 4,303 | 4,202 | 3,726 | 2,971 | 2,834 | 4,179 | 3,710 | 3,280 | 3,799 |
| 3 | 1,005 | 1,433 | 1,637 | 955 | 1,732 | 3,751 | 3,530 | 1,501 | 1,633 | 1,855 | 957 | 2,056 |
| 4 | 394 | 323 | 833 | 657 | 528 | 1,612 | 3,370 | 2,102 | 1,397 | 909 | 429 | 656 |
| 5 | 158 | 113 | 135 | 368 | 349 | 488 | 1,349 | 1,984 | 1,510 | 795 | 363 | 272 |
| 6 | 44 | 41 | 36 | 77 | 174 | 281 | 395 | 748 | 1,311 | 788 | 321 | 175 |
| 7 | 52 | 17 | 24 | 38 | 43 | 120 | 211 | 262 | 474 | 546 | 238 | 135 |
| 8 | 39 | 23 | 6 | 11 | 23 | 44 | 134 | 112 | 155 | 178 | 220 | 110 |
| $9+$ | 41 | 19 | 8 | 20 | 14 | 22 | 43 | 56 | 163 | 116 | 132 | 84 |
| Total | 5,478 | 5,484 | 7,542 | 20,165 | 13,496 | 16,377 | 18,262 | 12,781 | 17,173 | 19,326 | 13,003 | 11,220 |
| SSB | 807 | 697 | 942 | 817 | 897 | 1,637 | 2,174 | 1,874 | 1,545 | 1,216 | 1,035 | 1,082 |
| ('000t) |  |  |  |  |  |  |  |  |  |  |  |  |


| AGE <br> (RINGS) | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 6,198 | 9,416 | 4,449 | 5,087 | 24,735 | 6,837 | 23,055 | 9,829 | 5,183 | 3,114 | 6,823 |
| 2 | 4,557 | 6,363 | 5,747 | 3,078 | 2,922 | 12,290 | 4,875 | 18,949 | 3,415 | 2,055 | 3,772 |
| 3 | 2,824 | 3,287 | 2,520 | 4,725 | 2,156 | 3,083 | 8,220 | 3,081 | 9,191 | 3,649 | 1,997 |
| 4 | 1,087 | 1,696 | 1,625 | 1,116 | 3,139 | 1,462 | 1,390 | 4,189 | 2,167 | 5,790 | 2,098 |
| 5 | 311 | 692 | 982 | 506 | 1,006 | 1,676 | 795 | 675 | 2,590 | 1,213 | 4,175 |
| 6 | 99 | 259 | 445 | 314 | 483 | 450 | 1,031 | 495 | 317 | 1,175 | 618 |
| 7 | 83 | 79 | 170 | 139 | 266 | 170 | 244 | 568 | 328 | 140 | 562 |
| 8 | 133 | 78 | 45 | 54 | 120 | 98 | 121 | 146 | 342 | 127 | 84 |
| $9+$ | 206 | 158 | 121 | 87 | 97 | 59 | 150 | 178 | 186 | 107 | 70 |
| Total | 18,786 | 22,028 | 16,104 | 15,107 | 34,928 | 26,124 | 39,881 | 38,110 | 23,722 | 16,805 | 20,199 |
| SSB(‘000t) | 1,446 | 1,780 | 1,792 | 1,534 | 1,833 | 2,622 | 2,948 | 2,999 | 2,584 | 1,868 | 2,130 |
|  |  |  |  |  |  |  |  |  |  |  |  |

Table 2.3.2.1: North Sea autumn spawners. Fortnightly time periods sampled and survey effort in 2006/2007.

NL - Netherlands, FRG - Federal Republic of Germany

| Area | Time period | Samples available | Vessel days | Nation | Coverage |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Orkney/Shetland | 01-15 Sep. | 87 | 5 | GER | Total |
|  | 16-30 Sep. | 78 | 5 | GER | Total |
| Buchan | 01-15 Sep. | None |  |  |  |
|  | 16-30 Sep. | 78 | 5 | NL | Total |
| Central North | $01-15$ Sep. | None |  |  |  |
| Sea | 16-30 Sep. | 62 | 4 | NL | Total |
|  | 01-15 Oct. | None |  |  |  |
| Southern North | 16-31 Dec. | 77 | 4 | NL | Total |
| Sea | 01-15 Jan. | 104 | 7 | GER | Total |
|  | 16-31 Jan. | 82 | 5 | NL | Total |

Table 2.3.2.2: North Sea autumn spawners. Number of samples taken and sampling effort for the herring larvae surveys in Orkney/Shetland, Buchan, Central North Sea and Southern North Sea by year

| Year | Samples | Vessel-days (sampling) |
| :--- | :---: | :---: |
| $1988 / 89$ | 1355 | 98 |
| $1989 / 90$ | 1300 | 96 |
| $1990 / 91$ | 634 | 49 |
| $1991 / 92$ | 738 | 51 |
| $1992 / 93$ | 498 | 31 |
| $1993 / 94$ | 491 | 34 |
| $1994 / 95$ | 450 | 33 |
| $1995 / 96$ | 421 | 26 |
| $1996 / 97$ | 469 | 32 |
| $1997 / 98$ | 456 | 29 |
| $1998 / 99$ | 531 | 37 |
| $1999 / 00$ | 645 | 38 |
| $2000 / 01$ | 696 | 53 |
| $2001 / 02$ | 534 | 32 |
| $2002 / 03$ | 533 | 35 |
| $2003 / 04$ | 568 | 35 |
| $2004 / 05$ | 483 | 33 |
| $2005 / 06$ | 543 | 36 |
| $2006 / 07$ | 568 | 35 |

Table 2.3.2.3: North Sea autumn spawners. Estimated abundances of herring larvae $<\mathbf{1 0} \mathbf{~ m m}$ long ( $<11 \mathrm{~mm}$ for the SNS), by standard sampling area and time periods. The number of larvae are expressed as mean number per ICES rectangle * $10^{9}$

|  | Orkney/Shetland |  | Buchan |  | Central North Sea |  |  | Southern North Sea |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Period | $\begin{aligned} & 1-15 \\ & \text { Sep. } \end{aligned}$ | $\begin{gathered} 16-30 \\ \text { Sep. } \end{gathered}$ | $\begin{aligned} & 1-15 \\ & \text { Sep. } \end{aligned}$ | $\begin{gathered} 16-30 \\ \text { Sep. } \end{gathered}$ | $\begin{aligned} & 1-15 \\ & \text { Sep. } \end{aligned}$ | $\begin{gathered} 16-30 \\ \text { Sep. } \end{gathered}$ | $\begin{aligned} & 1-15 \\ & \text { Oct. } \end{aligned}$ | $\begin{aligned} & \text { 16-31 } \\ & \text { Dec. } \end{aligned}$ | $\begin{gathered} 1-15 \\ \text { Jan. } \end{gathered}$ | $\begin{gathered} \text { 16-31 } \\ \text { Jan. } \end{gathered}$ |
| 1972 | 1133 | 4583 | 30 |  | 165 | 88 | 134 | 2 | 46 |  |
| 1973 | 2029 | 822 | 3 | 4 | 492 | 830 | 1213 |  |  | 1 |
| 1974 | 758 | 421 | 101 | 284 | 81 |  | 1184 |  | 10 |  |
| 1975 | 371 | 50 | 312 |  |  | 90 | 77 | 1 | 2 |  |
| 1976 | 545 | 81 |  | 1 | 64 | 108 |  |  | 3 |  |
| 1977 | 1133 | 221 | 124 | 32 | 520 | 262 | 89 | 1 |  |  |
| 1978 | 3047 | 50 |  | 162 | 1406 | 81 | 269 | 33 | 3 |  |
| 1979 | 2882 | 2362 | 197 | 10 | 662 | 131 | 507 |  | 111 | 89 |
| 1980 | 3534 | 720 | 21 | 1 | 317 | 188 | 9 | 247 | 129 | 40 |
| 1981 | 3667 | 277 | 3 | 12 | 903 | 235 | 119 | 1456 |  | 70 |
| 1982 | 2353 | 1116 | 340 | 257 | 86 | 64 | 1077 | 710 | 275 | 54 |
| 1983 | 2579 | 812 | 3647 | 768 | 1459 | 281 | 63 | 71 | 243 | 58 |
| 1984 | 1795 | 1912 | 2327 | 1853 | 688 | 2404 | 824 | 523 | 185 | 39 |
| 1985 | 5632 | 3432 | 2521 | 1812 | 130 | 13039 | 1794 | 1851 | 407 | 38 |
| 1986 | 3529 | 1842 | 3278 | 341 | 1611 | 6112 | 188 | 780 | 123 | 18 |
| 1987 | 7409 | 1848 | 2551 | 670 | 799 | 4927 | 1992 | 934 | 297 | 146 |
| 1988 | 7538 | 8832 | 6812 | 5248 | 5533 | 3808 | 1960 | 1679 | 162 | 112 |
| 1989 | 11477 | 5725 | 5879 | 692 | 1442 | 5010 | 2364 | 1514 | 2120 | 512 |
| 1990 |  | 10144 | 4590 | 2045 | 19955 | 1239 | 975 | 2552 | 1204 |  |
| 1991 | 1021 | 2397 |  | 2032 | 4823 | 2110 | 1249 | 4400 | 873 |  |
| 1992 | 189 | 4917 |  | 822 | 10 | 165 | 163 | 176 | 1616 |  |
| 1993 |  | 66 |  | 174 |  | 685 | 85 | 1358 | 1103 |  |
| 1994 | 26 | 1179 |  |  |  | 1464 | 44 | 537 | 595 |  |
| 1995 |  | 8688 |  |  |  |  | 43 | 74 | 230 | 164 |
| 1996 |  | 809 |  | 184 |  | 564 |  | 337 | 675 | 691 |
| 1997 |  | 3611 |  | 23 |  |  |  | 9374 | 918 | 355 |
| 1998 |  | 8528 |  | 1490 | 205 | 66 |  | 1522 | 953 | 170 |
| 1999 |  | 4064 |  | 185 |  | 134 | 181 | 804 | 1260 | 344 |
| 2000 |  | 3352 | 28 | 83 |  | 376 |  | 7346 | 338 | 106 |
| 2001 |  | 11918 |  | 164 |  | 1604 |  | 971 | 5531 | 909 |
| 2002 |  | 6669 |  | 1038 |  |  | 3291 | 2008 | 260 | 925 |
| 2003 |  | 3199 |  | 2263 |  | 12018 | 3277 | 12048 | 3109 | 1116 |
| 2004 |  | 7055 |  | 3884 |  | 5545 |  | 7055 | 2052 | 4175 |
| 2005 |  | 3380 |  | 1364 |  | 5614 |  | 498 | 3999 | 4822 |
| 2006 | 6311 | 2312 |  | 280 |  | 2259 |  | 10858 | 2700 | 2106 |

Table 2.3.2.4: North Sea autumn spawners. Parameter estimates obtained on fitting the MLAI model to the estimates of larval abundance by area and time-period. Model fitted to abundances of larvae $<\mathbf{1 0} \mathbf{~ m m}$ in length ( $\mathbf{1 1} \mathbf{~ m m}$ for the southern North Sea).
a) Analysis of variance of the model fit

|  | DF | Sum <br> of Squares | Mean <br> Square | F Value | P |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Model | 44 | 169.97 | 3.863 | 8.37 | $<0.0001$ |
| Error | 240 | 110.81 | 0.462 |  |  |
| C Total | 284 | 280.79 |  |  |  |

b) Estimates of parameters

Reference Mean

| Estimate | Standard Error |  |
| :--- | :--- | :--- |
| 6.81331 | 0.5502 | Reference: 1972, Orkney/Shetland 09/01 - 09/15 |

Year Effects

| Year | Estimate | Standard Error | Year | Estimate | Standard Error |
| :--- | :---: | :---: | :---: | :---: | :---: |
| 1973 | 0.36421 | 0.68652 | 1990 | 2.93112 | 0.62998 |
| 1974 | -0.13565 | 0.73565 | 1991 | 2.28800 | 0.68255 |
| 1975 | -1.20382 | 0.74756 | 1992 | 1.52678 | 0.72159 |
| 1976 | -1.30869 | 0.73370 | 1993 | 1.18798 | 0.69815 |
| 1977 | -0.39908 | 0.70308 | 1994 | 0.81965 | 0.73606 |
| 1978 | -0.21089 | 0.71378 | 1995 | 0.91582 | 0.72536 |
| 1979 | 0.46102 | 0.68699 | 1996 | 1.59260 | 0.76399 |
| 1980 | 0.08025 | 0.68409 | 1997 | 1.83934 | 0.71657 |
| 1981 | 0.46853 | 0.68072 | 1998 | 2.10117 | 0.67352 |
| 1982 | 0.83775 | 0.61811 | 1999 | 1.91752 | 0.67715 |
| 1983 | 1.08677 | 0.63376 | 2000 | 1.50754 | 0.69242 |
| 1984 | 1.67991 | 0.61516 | 2001 | 2.66180 | 0.70501 |
| 1985 | 2.10061 | 0.59342 | 2002 | 2.50106 | 0.68429 |
| 1986 | 1.44595 | 0.61315 | 2003 | 3.39638 | 0.69663 |
| 1987 | 2.01122 | 0.60506 | 2004 | 3.56639 | 0.73851 |
| 1988 | 2.69650 | 0.59323 | 2005 | 3.05304 | 0.68755 |
| 1989 | 2.67140 | 0.60716 | 2006 | 2.56592 | 0.70976 |

Sampling Unit Effects

| Sampling Unit | Estimate | Standard Error |
| :--- | :---: | :---: |
| Or/Shet 16-30 Sep | -0.76580 | 0.31575 |
| Buchan 01-15 Sep | -1.79146 | 0.41538 |
| Buchan 16-30 Sep | -2.53508 | 0.34628 |
| CNS 01-15 Sep | -1.62701 | 0.40196 |
| CNS 16-30 Sep | -1.44483 | 0.34858 |
| CNS 01-15 Oct | -2.05251 | 0.37851 |
| CNS 16-31 Oct | -4.13775 | 0.52425 |
| SNS 12-31 Dec | -1.79694 | 0.37487 |
| SNS 01-15 Jan | -2.46444 | 0.32487 |
| SNS 16-31 Jan | -3.43855 | 0.36116 |

Table 2.3.2.5: North Sea autumn spawners. Time-series of the Multiplicative Larval Abundance Index (MLAI). The original MLAI is given in the second column. MLAI plus is the sum of the MLAI and the value of the reference area (Orkney/Shetlands, $1^{\text {st }}-15^{\text {th }}$ September 1972). This estimate is then unlogged (eMLAI) and divided by 100 ( MLAI $_{\text {assess }}$ ). The MLAI $_{\text {assess }}$ describes the time-series that is used in the assessment.

| Reference Value: | 6.81331 |
| :--- | ---: |


| Year | MLAI | MLAI ${ }_{\text {plus }}$ | eMLAI | $\mathrm{MLAI}_{\text {assess }}$ |
| :---: | :---: | :---: | :---: | :---: |
| 1973 | 0.36421 | 7.1775 | 1309.66 | 13.097 |
| 1974 | -0.13565 | 6.6777 | 794.46 | 7.945 |
| 1975 | -1.20382 | 5.6095 | 273 | 2.73 |
| 1976 | -1.30869 | 5.5046 | 245.82 | 2.458 |
| 1977 | -0.39908 | 6.4142 | 610.47 | 6.105 |
| 1978 | -0.21089 | 6.6024 | 736.88 | 7.369 |
| 1979 | 0.46102 | 7.2743 | 1442.78 | 14.428 |
| 1980 | 0.08025 | 6.8936 | 985.91 | 9.859 |
| 1981 | 0.46853 | 7.2818 | 1453.66 | 14.537 |
| 1982 | 0.83775 | 7.6511 | 2102.88 | 21.029 |
| 1983 | 1.08677 | 7.9001 | 2697.49 | 26.975 |
| 1984 | 1.67991 | 8.4932 | 4881.58 | 48.816 |
| 1985 | 2.10061 | 8.9139 | 7434.74 | 74.347 |
| 1986 | 1.44595 | 8.2593 | 3863.25 | 38.632 |
| 1987 | 2.01122 | 8.8245 | 6799.02 | 67.99 |
| 1988 | 2.6965 | 9.5098 | 13491.46 | 134.915 |
| 1989 | 2.6714 | 9.4847 | 13157.08 | 131.571 |
| 1990 | 2.93112 | 9.7444 | 17058.94 | 170.589 |
| 1991 | 2.288 | 9.1013 | 8967.01 | 89.67 |
| 1992 | 1.52678 | 8.3401 | 4188.46 | 41.885 |
| 1993 | 1.18798 | 8.0013 | 2984.82 | 29.848 |
| 1994 | 0.81965 | 7.633 | 2065.17 | 20.652 |
| 1995 | 0.91582 | 7.7291 | 2273.63 | 22.736 |
| 1996 | 1.5926 | 8.4059 | 4473.42 | 44.734 |
| 1997 | 1.83934 | 8.6527 | 5725.33 | 57.253 |
| 1998 | 2.10117 | 8.9145 | 7438.89 | 74.389 |
| 1999 | 1.91752 | 8.7308 | 6190.9 | 61.909 |
| 2000 | 1.50754 | 8.3208 | 4108.65 | 41.087 |
| 2001 | 2.6618 | 9.4751 | 13031.31 | 130.313 |
| 2002 | 2.50106 | 9.3144 | 11096.36 | 110.964 |
| 2003 | 3.39638 | 10.2097 | 27165.27 | 271.653 |
| 2004 | 3.56639 | 10.3797 | 32199.47 | 321.995 |
| 2005 | 3.05304 | 9.8663 | 19270.81 | 192.708 |
| 2006 | 2.56592 | 9.3792 | 11839.94 | 118.399 |

Table 2.3.3.1. North Sea herring. Indices of 2-5+ ringers from the $\mathbf{1}^{\text {st }}$ quarter IBTS

| Year of SAmpling | 2-RINGER | 3-RINGER | 4-RINGER | 5+ RINGER |
| :---: | :---: | :---: | :---: | :---: |
| 1983 | 139 | 45 | 14 | 24 |
| 1984 | 161 | 61 | 27 | 10 |
| 1985 | 722 | 282 | 42 | 28 |
| 1986 | 782 | 276 | 79 | 28 |
| 1987 | 918 | 116 | 59 | 49 |
| 1988 | 4163 | 792 | 58 | 25 |
| 1989 | 875 | 339 | 89 | 9 |
| 1990 | 462 | 280 | 269 | 71 |
| 1991 | 693 | 259 | 222 | 146 |
| 1992 | 437 | 193 | 55 | 92 |
| 1993 | 787 | 223 | 45 | 66 |
| 1994 | 1167 | 213 | 69 | 43 |
| 1995 | 1393 | 279 | 37 | 7 |
| 1996 | 198 | 33 | 10 | 8 |
| 1997 | 507 | 163 | 31 | 20 |
| 1998 | 792 | 96 | 21 | 18 |
| 1999 | 451 | 501 | 98 | 36 |
| 2000 | 199 | 155 | 59 | 9 |
| 2001 | 1129 | 317 | 94 | 68 |
| 2002 | 658 | 338 | 25 | 20 |
| 2003 | 1556 | 612 | 360 | 53 |
| 2004 | 451 | 777 | 112 | 171 |
| 2005 | 214 | 356 | 389 | 131 |
| 2006 | 1464 | 330 | 252 | 339 |
| 2007 | 41 | 18 | 8 | 41 |

Table 2.3.3.2. North Sea herring. Estimates of mean number per hour per statistical rectangle from $1^{\text {st }}$ quarter IBTS 2007. Means for age groups in "Roundfish areas" (*) and in all areas. In the index 2-5+ for all areas, the findings in RF8 and RF9 are not included.

| Area | Total | Mean per statistical rectangle |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | AGE GROUP (WR) |  |  |  |  |
|  |  | 1 | 2 | 3 | 4 | 5+ |
| All areas |  | 1336 | 41 | 18 | 8 | 41 |
| RF1 | 294.9 | 0.0 | 33.4 | 64.0 | 28.2 | 169.3 |
| RF2 | 158.8 | 149.5 | 8.3 | 0.8 | 0.0 | 0.2 |
| RF3 | 95.8 | 84.9 | 9.8 | 0.8 | 0.0 | 0.4 |
| RF4 | 83.1 | 36.9 | 23.6 | 7.4 | 14.9 | 0.3 |
| RF5 | 645.7 | 601.5 | 24.2 | 13.4 | 3.8 | 2.7 |
| RF6 | 2957.9 | 2912.5 | 44.7 | 0.4 | 0.0 | 0.2 |
| RF7 | 4755.0 | 4566.0 | 186.8 | 1.9 | 0.0 | 0.3 |
| RF8 | 1005.2 | 764.7 | 213.5 | 10.6 | 6.9 | 9.5 |
| RF9 | 13332.9 | 10385.5 | 2636.4 | 266.3 | 44.7 | 0.0 |

[^4]Table 2.3.3.3. North Sea herring. Indices of 1-ringers from the IBTS $\mathbf{1}^{\text {st }}$ Quarter. Estimation of the small sized component (possibly Downs herring) in different areas. " North Sea" = total area of sampling minus IIIa.

| Year <br> class | Year <br> of sampling | All <br> 1-ringers <br> in total area (no/hour) | Small $<13 \mathrm{~cm}$ <br> 1-ringers <br> in total area <br> (no/hour) | Proportion of small in total area vs. all sizes | Small $<13 \mathrm{~cm}$ <br> 1-ringers <br> in North Sea <br> (no/hour) | Proportion of small in North Sea vs. all sizes | Proportion of small in IIIa vs small in total area |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 1979 | 168 | 11 | 0.07 | 12 | 0.07 | 0 |
| 1978 | 1980 | 316 | 108 | 0.34 | 106 | 0.34 | 0.09 |
| 1979 | 1981 | 495 | 51 | 0.1 | 41 | 0.08 | 0.25 |
| 1980 | 1982 | 798 | 177 | 0.22 | 185 | 0.23 | 0.03 |
| 1981 | 1983 | 1270 | 192 | 0.15 | 185 | 0.15 | 0.10 |
| 1982 | 1984 | 1516 | 346 | 0.23 | 297 | 0.20 | 0.20 |
| 1983 | 1985 | 2097 | 315 | 0.15 | 298 | 0.14 | 0.12 |
| 1984 | 1986 | 2663 | 596 | 0.22 | 390 | 0.15 | 0.39 |
| 1985 | 1987 | 3693 | 628 | 0.17 | 529 | 0.14 | 0.22 |
| 1986 | 1988 | 4394 | 2371 | 0.54 | 720 | 0.16 | 0.72 |
| 1987 | 1989 | 2332 | 596 | 0.26 | 531 | 0.23 | 0.17 |
| 1988 | 1990 | 1062 | 70 | 0.07 | 62 | 0.06 | 0.18 |
| 1989 | 1991 | 1287 | 330 | 0.26 | 337 | 0.26 | 0.05 |
| 1990 | 1992 | 1268 | 125 | 0.1 | 130 | 0.10 | 0.03 |
| 1991 | 1993 | 2794 | 676 | 0.24 | 176 | 0.06 | 0.76 |
| 1992 | 1994 | 1752 | 283 | 0.16 | 240 | 0.14 | 0.21 |
| 1993 | 1995 | 1346 | 449 | 0.33 | 445 | 0.33 | 0.08 |
| 1994 | 1996 | 1891 | 604 | 0.32 | 467 | 0.25 | 0.28 |
| 1995 | 1997 | 4405 | 1356 | 0.31 | 1089 | 0.25 | 0.25 |
| 1996 | 1998 | 2276 | 1322 | 0.58 | 1399 | 0.61 | 0.02 |
| 1997 | 1999 | 753 | 152 | 0.2 | 149 | 0.20 | 0.09 |
| 1998 | 2000 | 3725 | 1117 | 0.3 | 991 | 0.27 | 0.18 |
| 1999 | 2001 | 2499 | 328 | 0.13 | 307 | 0.12 | 0.13 |
| 2000 | 2002 | 4065 | 1553 | 0.38 | 1471 | 0.36 | 0.12 |
| 2001 | 2003 | 2765 | 717 | 0.26 | 237 | 0.09 | 0.69 |
| 2002 | 2004 | 979 | 665 | 0.68 | 710 | 0.73 | 0.01 |
| 2003 | 2005 | 1002 | 340 | 0.34 | 356 | 0.36 | 0.03 |
| 2004 | 2006 | 922 | 122 | 0.13 | 128 | 0.14 | 0.02 |
| 2005 | 2007 | 1336 | 304 | 0.23 | 305 | 0.23 | 0.07 |

Table 2.3.3.4 North Sea herring. Density and abundance estimates of 0-ringers caught in February during the IBTS. Values given for year classes by areas are density estimates in numbers per square metre. Total abundance is found by multiplying density by area and summing up.

| Area | NORTH west | NORTH EAST | Central WEST | Central EAST | South WEST | South <br> EAST | DIv. IIIA | South’ BIGHT | 0-RINGER ABUNDANCE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { Area } \mathrm{m}^{2} \mathrm{x} \\ & 10^{9} \end{aligned}$ | 83 | 34 | 86 | 102 | 37 | 93 | 31 | 31 |  |
| Year class |  |  |  |  |  |  |  |  | no. in $10^{9}$ |
| 1976 | 0.054 | 0.014 | 0.122 | 0.005 | 0.008 | 0.002 | 0.002 | 0.016 | 17.1 |
| 1977 | 0.024 | 0.024 | 0.05 | 0.015 | 0.056 | 0.013 | 0.006 | 0.034 | 13.1 |
| 1978 | 0.176 | 0.031 | 0.061 | 0.02 | 0.01 | 0.005 | 0.074 | 0 | 52.1 |
| 1979 | 0.061 | 0.195 | 0.262 | 0.408 | 0.226 | 0.143 | 0.099 | 0.053 | 101.1 |
| 1980 | 0.052 | 0.001 | 0.145 | 0.115 | 0.089 | 0.339 | 0.248 | 0.187 | 76.7 |
| 1981 | 0.197 | 0 | 0.289 | 0.199 | 0.215 | 0.645 | 0.109 | 0.036 | 133.9 |
| 1982 | 0.025 | 0.011 | 0.068 | 0.248 | 0.29 | 0.309 | 0.47 | 0.14 | 91.8 |
| 1983 | 0.019 | 0.007 | 0.114 | 0.268 | 0.271 | 0.473 | 0.339 | 0.377 | 115 |
| 1984 | 0.083 | 0.019 | 0.303 | 0.259 | 0.996 | 0.718 | 0.277 | 0.298 | 181.3 |
| 1985 | 0.116 | 0.057 | 0.421 | 0.344 | 0.464 | 0.777 | 0.085 | 0.084 | 177.4 |
| 1986 | 0.317 | 0.029 | 0.73 | 0.557 | 0.83 | 0.933 | 0.048 | 0.244 | 270.9 |
| 1987 | 0.078 | 0.031 | 0.417 | 0.314 | 0.159 | 0.618 | 0.483 | 0.495 | 168.9 |
| 1988 | 0.036 | 0.02 | 0.095 | 0.096 | 0.151 | 0.411 | 0.181 | 0.016 | 71.4 |
| 1989 | 0.083 | 0.03 | 0.04 | 0.094 | 0.013 | 0.035 | 0.041 | 0 | 25.9 |
| 1990 | 0.075 | 0.053 | 0.202 | 0.158 | 0.121 | 0.198 | 0.086 | 0.196 | 69.9 |
| 1991 | 0.255 | 0.39 | 0.431 | 0.539 | 0.5 | 0.369 | 0.298 | 0.395 | 200.7 |
| 1992 | 0.168 | 0.039 | 0.672 | 0.444 | 0.734 | 0.268 | 0.345 | 0.285 | 190.1 |
| 1993 | 0.358 | 0.212 | 0.26 | 0.187 | 0.12 | 0.119 | 0.223 | 0.028 | 101.7 |
| 1994 | 0.148 | 0.024 | 0.417 | 0.381 | 0.332 | 0.148 | 0.252 | 0.169 | 126.9 |
| 1995 | 0.26 | 0.086 | 0.699 | 0.092 | 0.266 | 0.018 | 0.001 | 0.02 | 106.2 |
| 1996 | 0.003 | 0.004 | 0.935 | 0.135 | 0.436 | 0.379 | 0.039 | 0.032 | 148.1 |
| 1997 | 0.042 | 0.021 | 0.338 | 0.064 | 0.178 | 0.035 | 0.023 | 0.083 | 53.1 |
| 1998 | 0.1 | 0.056 | 1.15 | 0.592 | 0.998 | 0.265 | 0.28 | 0.127 | 244.0 |
| 1999 | 0.045 | 0.011 | 0.799 | 0.2 | 0.514 | 0.22 | 0.107 | 0.026 | 137.1 |
| 2000 | 0.284 | 0.011 | 1.052 | 0.197 | 1.156 | 0.376 | 0.063 | 0.006 | 214.8 |
| 2001 | 0.08 | 0.019 | 0.566 | 0.473 | 0.567 | 0.247 | 0.209 | 0.226 | 161.8 |
| 2002 | 0.141 | 0.04 | 0.287 | 0.028 | 0.121 | 0.045 | 0.003 | 0.157 | 54.4 |
| 2003 | 0.045 | 0.005 | 0.284 | 0.074 | 0.106 | 0.021 | 0.022 | 0.154 | 47.3 |
| 2004 | 0.017 | 0.010 | 0.189 | 0.089 | 0.268 | 0.187 | 0.027 | 0.198 | 61.3 |
| 2005 | 0.013 | 0.018 | 0.327 | 0.081 | 0.633 | 0.184 | 0.007 | 0.131 | 83.1 |
| 2006 | 0.004 | 0.001 | 0.240 | 0.025 | 0.098 | 0.018 | 0.040 | 0.228 | 37.2 |

## Table 2.4.1.1: North Sea Herring: Mean weight-at-age (wr) in the third quarter, in Divisions IVa, IVb and IIIa

| RING | Third quarter mean wts in Catch (Divisions iva, IVb \& IIIA) |  |  |  |  |  |  |  |  |  |  | July acoustic surver |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 75 | 43 | 54 | 62 | 54 | 69 | 50 | 65 | 45 | 53 | 61 | 45 | 45 | 52 | 52 | 46 | 50 | 45 | 46 | 35 | 43 | 45 |
| 2 | 135.1 | 129 | 131 | 128 | 123 | 136 | 140 | 119 | 125 | 124 | 139 | 119 | 120 | 109 | 118 | 118 | 127 | 138 | 104 | 116 | 135 | 127 |
| 3 | 186.3 | 175 | 172 | 163 | 172 | 167 | 177 | 177 | 159 | 177 | 163 | 196 | 168 | 198 | 171 | 180 | 162 | 172 | 185 | 139 | 171 | 158 |
| 4 | 224.3 | 220 | 209 | 193 | 201 | 199 | 200 | 198 | 203 | 201 | 192 | 253 | 233 | 238 | 207 | 218 | 204 | 194 | 209 | 206 | 181 | 188 |
| 5 | 229.3 | 247 | 237 | 228 | 228 | 218 | 224 | 210 | 234 | 234 | 205 | 262 | 256 | 275 | 236 | 232 | 228 | 224 | 214 | 231 | 229 | 188 |
| 6 | 252.6 | 255 | 263 | 252 | 241 | 237 | 244 | 236 | 250 | 249 | 242 | 299 | 245 | 307 | 267 | 261 | 237 | 247 | 243 | 253 | 248 | 225 |
| 7 | 291.6 | 278 | 269 | 263 | 266 | 262 | 252 | 247 | 264 | 261 | 257 | 306 | 265 | 289 | 272 | 295 | 255 | 261 | 281 | 262 | 253 | 243 |
| 8 | 300.3 | 295 | 313 | 275 | 286 | 288 | 281 | 272 | 262 | 287 | 260 | 325 | 269 | 308 | 230 | 300 | 286 | 280 | 290 | 279 | 274 | 244 |
| 9+ | 302.3 | 295 | 298 | 306 | 271 | 298 | 298 | 282 | 299 | 270 | 285 | 335 | 329 | 363 | 260 | 280 | 294 | 249 | 307 | 270 | 295 | 265 |

Weights-at-age in the catch for 1996 to 2001 were revised by SG Rednose for details of the revision see last years report (ICES ACFM).

Table 2.4.2.1 North Sea herring. Percentage maturity at 2-, 3- and 4+ ring for Autumn Spawning herring in the North Sea. The values are derived from the acoustic survey for 1988 to 2006.

| Year $\backslash$ Ring | 2 | 3 | $>3$ |
| :--- | :--- | :--- | :--- |
| 1988 | 65.6 | 87.7 | 100 |
| 1989 | 78.7 | 93.9 | 100 |
| 1990 | 72.6 | 97.0 | 100 |
| 1991 | 63.8 | 98.0 | 100 |
| 1992 | 51.3 | 100 | 100 |
| 1993 | 47.1 | 62.9 | 100 |
| 1994 | 72.1 | 85.8 | 100 |
| 1995 | 72.6 | 95.4 | 100 |
| 1996 | 60.5 | 97.5 | 100 |
| 1997 | 64.0 | 94.2 | 100 |
| 1998 | 64.0 | 89.0 | 100 |
| 1999 | 81.0 | 91.0 | 100 |
| 2000 | 66.0 | 96.0 | 100 |
| 2001 | 77.0 | 92.0 | 100 |
| 2002 | 86.0 | 97.0 | 100 |
| 2003 | 43.0 | 93.0 | 100 |
| 2004 | 69.8 | 64.9 | 100 |
| 2005 | 76.0 | 97.0 | 100 |
| 2006 | 66.0 | 88.0 | 100 |

Table 2.6.1 North Sea herring. Years of duration of survey and years used in the assessment.

| SURVEY | Age range | Years survey has <br> BEEN RUNNING | Years used in <br> ASSESSMENT |
| :--- | :--- | :---: | :---: |
| MLAI (Larvae survey) | SSB | $1972-2006$ | $1973-2006$ |
| IBTS 1 ${ }^{\text {st }}$ Quarter (Trawl survey) | $1-5 \mathrm{wr}$ | $1971-2007$ | $1984-2007$ |
| Acoustic (+trawl) | 1 wr | $1995-2006$ | $1997-2006$ |
|  | $2-9+\mathrm{wr}$ | $1984-2006$ | $1989-2006$ |
| MIK net | 0 wr | $1977-2007$ | $1992-2007$ |

Table 2.6.2 North Sea herring. Percentage change in estimated mean F2-6, SSB, TSB and Recruitment in years 2001 to 2006 produced by removing points values in the surveys that show high residuals in the assessment.

| Year | Acoust IC 1 WR IN 2005 AND 20066 | IBTS <br> 2 WR IN 2006 AND 2007 | Mik est <br> 0 WR IN <br> 2004 | $\begin{gathered} \text { MLAI } \\ \text { EST OF SSB } \\ \text { 2003 AND } 2004 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: |
|  | F 2-6 |  |  |  |
| 2001 | 0\% | 0\% | 0\% | 2\% |
| 2002 | 0\% | 0\% | 0\% | 3\% |
| 2003 | 0\% | 0\% | 0\% | 4\% |
| 2004 | -1\% | 1\% | -1\% | 4\% |
| 2005 | -1\% | 1\% | -1\% | 4\% |
| 2006 | -3\% | 2\% | -3\% | 5\% |
|  | SSB |  |  |  |
| 2001 | 0\% | 0\% | 0\% | -3\% |
| 2002 | 0\% | 0\% | 0\% | -3\% |
| 2003 | 1\% | 0\% | 1\% | -4\% |
| 2004 | 1\% | -1\% | 1\% | -5\% |
| 2005 | 1\% | -1\% | 1\% | -5\% |
| 2006 | 3\% | -2\% | 3\% | -6\% |
|  | TSB |  |  |  |
| 2001 | 0\% | 0\% | 0\% | -2\% |
| 2002 | 1\% | -1\% | 1\% | -3\% |
| 2003 | 1\% | 0\% | 1\% | -3\% |
| 2004 | 1\% | -1\% | 1\% | -4\% |
| 2005 | 1\% | 0\% | 2\% | -4\% |
| 2006 | 0\% | 2\% | 2\% | -4\% |
| Recruitment (yearclass) |  |  |  |  |
| 2000 | 1\% | -1\% | 1\% | -3\% |
| 2001 | 1\% | -1\% | 1\% | -2\% |
| 2002 | 2\% | -1\% | 1\% | -2\% |
| 2003 | 8\% | -7\% | 7\% | -2\% |
| 2004 | -19\% | 33\% | 0\% | -1\% |
| 2005 | -1\% | 1\% | -1\% | -1\% |

## Table 2.6.3 North Sea herring. Final assessment ICA log file . Note age=ringer.

Integrated Catch at Age Analysis
------------------------------
Version 1.4 w
K.R.Patterson

Fisheries Research Services
Marine Laboratory
Aberdeen
24 August 1999
Type * to change language
Enter the name of the index file -->index.txt
canum.txt
weca.txt
Stock weights in 2007 used for the year 2006
west.txt
Natural mortality in 2007 used for the year 2006
natmor.txt
Maturity ogive in 2007 used for the year 2006
matprop.txt
Name of age-structured index file (Enter if none) : -->fleet.txt
Name of the SSB index file (Enter if none) -->SSB
File not found: SSB
Name of the SSB index file (Enter if none) -->SSB.txt
No of years for separable constraint ?--> 5
Reference age for separable constraint ?--> 4
Constant selection pattern model (Y/N) ?-->y
$S$ to be fixed on last age ?--> 1.000000000000000
First age for calculation of reference $F$ ?--> 2
Last age for calculation of reference F ?--> 6
Use default weighting (Y/N) ?-->n
Enter relative weights at age
Weight for age 0--> 0.100000000000000
Weight for age 1--> 0.100000000000000
Weight for age 2--> 3.670000000000000
Weight for age 3--> 2.870000000000000
Weight for age 4--> 2.230000000000000
Weight for age 5--> 1.740000000000000
Weight for age 6--> 1.370000000000000
Weight for age 7--> 1.040000000000000
Weight for age 8--> 0.940000000000000
Weight for age 9--> 0.910000000000000
Enter relative weights by year
Weight for year 2002--> 1.000000000000000
Weight for year 2003--> 1.000000000000000
Weight for year 2004--> 1.000000000000000
Weight for year 2005--> 1.000000000000000
Weight for year 2006--> 1.000000000000000
Enter new weights for specified years and ages if needed
Enter year, age, new weight or $-1,-1,-1$ to end. -1 -1 -1.0000000000000000
Is the last age of Acoustic survey 1-9+ wr a plus-group (Y/N) ?-->y
Is the last age of IBTS1: 1-5+ wr a plus-group (Y/N) ?-->y
Is the last age of MIK $0-w r$ a plus-group ( $\mathrm{Y} / \mathrm{N}$ ) ?-->n

Table 2.6.3(cont) North Sea herring. Final assessment ICA log file . Note age=ringer.
You must choose a catchability model for each index.
Models: A Absolute: Index = Abundance . e
L Linear: Index $=$ Q. Abundance . e
$P$ Power: Index $=$ Q. Abundance^ K .e
where Q and K are parameters to be estimated, and e is a lognormally-distributed error.

Model for MLAI is to be A/L/P ?-->p
Model for Acoustic survey 1-9+ wr is to be A/L/P ?-->L
Model for IBTS1: $1-5+w r$ is to be $A / L / P$ ?-->L
Model for MIK 0-wr is to be A/L/P ?-->L
Fit a stock-recruit relationship (Y/N) ?-->y
Enter the time lag in years between spawning and the stock size
of fish aged 0 years on 1 January.
This will probably be 0 unless the stock is an autumn-spawning herring
in which case it will probably be 1 years.
Enter the lag in years (rounded up)--> 1
Enter lowest feasible F--> 2.0000000000000000E-02
Enter highest feasible F--> 0.500000000000000
Mapping the F-dimension of the SSQ surface

| F | SSQ |
| :---: | :---: |
| 0.02 | 123.3974328218 |
| 0.05 | 81.0648173194 |
| 0.07 | 60.9889192791 |
| 0.10 | 48.6616599867 |
| 0.12 | 40.4041715837 |
| 0.15 | 34.6450957408 |
| 0.17 | 30.5385725459 |
| 0.20 | 27.5729749766 |
| 0.22 | 25.4183334268 |
| 0.25 | 23.8529324093 |
| 0.27 | 22.7235141854 |
| 0.30 | 21.9217861351 |
| 0.32 | 21.3698238771 |
| 0.35 | 21.0106665969 |
| 0.37 | 20.8020895893 |
| 0.40 | 20.7123931302 |
| 0.42 | 20.7175014351 |
| 0.45 | 20.7989339979 |
| 0.47 | 20.9423659141 |
| 0.50 | 21.1365925135 |
|  |  |

No of years for separable analysis : 5
Age range in the analysis : 0 . . . 9
Year range in the analysis : 1960 . . . 2006
Number of indices of SSB : 1
Number of age-structured indices : 3
Stock-recruit relationship to be fitted.
Parameters to estimate : 45
Number of observations : 415

Table 2.6.3(cont) North Sea herring. Final assessment ICA log file . Note age=ringer.
Conventional single selection vector model to be fitted.


Table 2.6.4 North Sea herring. Final model fit ICA output. Note age=ringer

| Catch in Number |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 |
| 0 | 195. | 1269. | 142. | 443. | 497. | 157. | 375. | 645. |
| 1 | 2393. | 336. | 2147. | 1262. | 2972. | 3209. | 1383. | 1674. |
| 2 | 1142. | 1889. | 270. | 2961. | 1548. | 2218. | 2570. | 1172. |
| 3 | 1967. | 480. | 797. | 177. | 2243. | 1325. | 741. | 1365. |
| 4 | 166. | 1456. | 335. | 158. | 148. | 2039. | 450. | 372. |
| 5 | 168. | 124. | 1082. | 81. | 149. | 145. | 890. | 298. |
| 6 | 113. | 158. | 127. | 230. | 95. | 152. | 45. | 393. |
| 7 | 126. | 61. | 145. | 22. | 256. | 118. | 65. | 68. |
| 8 | 129. | 56. | 86. | 42. | 26. | 413. | 96. | 82. |
| 9 | 142. | 88. | 87. | 51. | 58. | 78. | 236. | 173. |


| Catch in Number |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 |
| 0 | 839. | 112. | 898. | 684. | 750. | 289. | 996. | 264. |
| 1 | 2425. | 2503. | 1196. | 4379. | 3341. | 2368. | 846. | 2461. |
| 2 | 1795. | 1883. | 2003. | 1147. | 1441. | 1344. | 773. | 542. |
| 3 | 1494. | 296. | 884. | 663. | 344. | 659. | 362. | 260. |
| 4 | 621. | 133. | 125. | 208. | 131. | 150. | 126. | 141. |
| 5 | 157. | 191. | 50. | 27. | 33. | 59. | 56. | 57. |
| 6 | 145. | 50. | 61. | 31. | 5. | 31. | 22. | 16. |
| 7 | 163. | 43. | 8. | 27. | 0. | 4. | 5. | 9. |
| 8 | 14. | 27. | 12. | 0. | 1. | 1. | 2. | 3. |
| 9 | 92. | 25. | 12. | 12. | 0. | 1. | 1. | 1. |

Catch in Number

| AGE | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 238. | 257. | 130. | 542. | 1263. | 9520. | 11957. | 13297. |
| 1 | 127. | 144. | 169. | 159. | 245. | 872. | 1116. | 2449. |
| 2 | 902. | 45. | 5. | 34. | 134. | 284. | 299. | 574. |
| 3 | 117. | 186. | 6. | 10. | 92. | 57. | 230. | 216. |
| 4 | 52. | 11. | 5. | 10. | 32. | 40. | 34. | 105. |
| 5 | 35. | 7. | 0. | 2. | 22. | 29. | 14. | 26. |
| 6 | 6. | 4. | 0. | 0. | 2. | 23. | 7. | 23. |
| 7 | 4. | 2. | 0. | 1. | 1. | 19. | 8. | 13. |
| 8 | 1. | 1. | 0. | 1. | 0. | 6. | 4. | 11. |
| 9 | 0. | 0. | 0. | 0. | 0. | 1. | 1. | 12. |

Catch in Number

| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 6973. | 4211. | 3725. | 8229. | 3165. | 3058. | 1303. | 2387. |
| 1 | 1818. | 3253. | 4801. | 6836. | 7867. | 3146. | 3020. | 2139. |
| 2 | 1146. | 1326. | 1267. | 2137. | 2233. | 1594. | 899. | 1133. |
| 3 | 441. | 1182. | 841. | 668. | 1091. | 1364. | 779. | 557. |
| 4 | 202. | 369. | 466. | 467. | 384. | 809. | 861. | 549. |
| 5 | 81. | 125. | 130. | 246. | 256. | 212. | 388. | 501. |
| 6 | 23. | 44. | 62. | 75. | 128. | 124. | 80. | 205. |
| 7 | 25. | 20. | 21. | 24. | 38. | 61. | 54. | 39. |
| 8 | 11. | 13. | 14. | 8. | 15. | 20. | 29. | 26. |
| 9 | 19. | 16. | 15. | 8. | 9. | 9. | 12. | 13. |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Catch in Number |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 10331. | 10265. | 4499. | 7438. | 2311. | 431. | 260. | 1566. |
| 1 | 2303. | 3827. | 1785. | 1665. | 1606. | 480. | 978. | 304. |
| 2 | 1285. | 1176. | 1783. | 1444. | 642. | 688. | 1220. | 616. |
| 3 | 443. | 609. | 489. | 817. | 526. | 447. | 538. | 1059. |
| 4 | 362. | 306. | 348. | 232. | 172. | 285. | 276. | 294. |
| 5 | 361. | 216. | 109. | 119. | 58. | 109. | 176. | 136. |
| 6 | 376. | 226. | 92. | 55. | 23. | 31. | 89. | 69. |
| 7 | 152. | 188. | 76. | 41. | 9. | 12. | 15. | 28. |
| 8 | 39. | 87. | 70. | 69. | 17. | 19. | 17. | 10. |
| 9 | 23. | 42. | 47. | 29. | 4. | 6. | 4. | 2. |


| Catch in Number |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 0 | 1105. | 1833. | 730. | 369. | 716. | 1016. | 879. |
| 1 | 1172. | 614. | 838. | 617. | 207. | 716. | 222. |
| 2 | 623. | 843. | 580. | 1222. | 448. | 355. | 401. |
| 3 | 463. | 486. | 971. | 529. | 1366. | 486. | 311. |
| 4 | 647. | 279. | 292. | 836. | 543. | 1319. | 465. |
| 5 | 213. | 322. | 141. | 245. | 753. | 480. | 998. |
| 6 | 82. | 91. | 175. | 108. | 169. | 576. | 252. |
| 7 | 36. | 38. | 49. | 123. | 105. | 115. | 247. |
| 8 | 15. | 18. | 35. | 38. | 65. | 88. | 63. |
| 9 | 2. | 3. | 9. | 9. | 32. | 58. | 4. |


| Predicted Catch in Number |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2002 | 2003 | 2004 | 2005 | 2006 |
| 0 | 844.9 | 508.3 | 699.6 | 684.5 | 1034.3 |
| 1 | 1310.1 | 453.0 | 302.6 | 459.9 | 335.5 |
| 2 | 535.3 | 1268.2 | 484.4 | 352.6 | 398.9 |
| 3 | 906.2 | 524.4 | 1365.7 | 560.5 | 301.1 |
| 4 | 291.4 | 888.5 | 561.7 | 1545.3 | 464.3 |
| 5 | 174.9 | 217.5 | 722.7 | 477.7 | 948.9 |
| 6 | 170.4 | 118.0 | 159.8 | 553.5 | 262.8 |
| 7 | 52.9 | 111.0 | 83.7 | 118.4 | 294.3 |
| 8 | 33.0 | 33.1 | 75.8 | 59.9 | 60.8 |


| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 |
| 0 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 |
| 1 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 |
| 2 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.12600 |
| 3 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.17600 |
| 4 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.21100 |
| 5 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.24300 |
| 6 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 |
| 7 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 |
| 8 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 |
| 9 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 |
| 0 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 |
| 1 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 |
| 2 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.12600 |
| 3 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.17600 |
| 4 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.21100 |
| 5 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.24300 |
| 6 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 |
| 7 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 |
| 8 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 |
| 9 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 |


| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| 0 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.00700 | 0.01000 | 0.01000 |
| 1 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.04900 | 0.05900 | 0.05900 |
| 2 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.12600 | 0.11800 | 0.11800 | 0.11800 |
| 3 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.17600 | 0.14200 | 0.14900 | 0.14900 |
| 4 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.21100 | 0.18900 | 0.17900 | 0.17900 |
| 5 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.24300 | 0.21100 | 0.21700 | 0.21700 |
| 6 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.22200 | 0.23800 | 0.23800 |
| 7 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26700 | 0.26500 | 0.26500 |
| 8 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27400 | 0.27400 |
| 9 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27100 | 0.27500 | 0.27500 |


| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 0 | 0.01000 | 0.00900 | 0.00600 | 0.01100 | 0.01100 | 0.01700 | 0.01900 | 0.01700 |
| 1 | 0.05900 | 0.03600 | 0.06700 | 0.03500 | 0.05500 | 0.04300 | 0.05500 | 0.05800 |
| 2 | 0.11800 | 0.12800 | 0.12100 | 0.09900 | 0.11100 | 0.11500 | 0.11400 | 0.13000 |
| 3 | 0.14900 | 0.16400 | 0.15300 | 0.15000 | 0.14500 | 0.15300 | 0.14900 | 0.16600 |
| 4 | 0.17900 | 0.19400 | 0.18200 | 0.18000 | 0.17400 | 0.17300 | 0.17700 | 0.18400 |
| 5 | 0.21700 | 0.21100 | 0.20800 | 0.21100 | 0.19700 | 0.20800 | 0.19300 | 0.20300 |
| 6 | 0.23800 | 0.22000 | 0.22100 | 0.23400 | 0.21600 | 0.23100 | 0.22900 | 0.21700 |
| 7 | 0.26500 | 0.25800 | 0.23800 | 0.25800 | 0.23700 | 0.24700 | 0.23600 | 0.23500 |
| 8 | 0.27400 | 0.27000 | 0.25200 | 0.27700 | 0.25300 | 0.26500 | 0.25000 | 0.25900 |
| 9 | 0.27500 | 0.29200 | 0.26200 | 0.29900 | 0.26300 | 0.25900 | 0.28700 | 0.27100 |


| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 0.01000 | 0.01000 | 0.00600 | 0.00900 | 0.01500 | 0.01500 | 0.02100 | 0.00900 |
| 1 | 0.05300 | 0.03300 | 0.05600 | 0.04200 | 0.01800 | 0.04400 | 0.05100 | 0.04500 |
| 2 | 0.10200 | 0.11500 | 0.13000 | 0.13000 | 0.11200 | 0.10800 | 0.11400 | 0.11500 |
| 3 | 0.17500 | 0.14500 | 0.15900 | 0.16900 | 0.15600 | 0.14800 | 0.14500 | 0.15100 |
| 4 | 0.18900 | 0.18900 | 0.18100 | 0.19800 | 0.18800 | 0.19500 | 0.18300 | 0.17100 |
| 5 | 0.20700 | 0.20400 | 0.21400 | 0.20700 | 0.20400 | 0.22700 | 0.21900 | 0.20700 |
| 6 | 0.22300 | 0.22800 | 0.24000 | 0.24300 | 0.21200 | 0.22600 | 0.23800 | 0.23300 |
| 7 | 0.23700 | 0.24400 | 0.25500 | 0.24700 | 0.26100 | 0.23500 | 0.24700 | 0.24500 |
| 8 | 0.24900 | 0.25600 | 0.27300 | 0.28300 | 0.28000 | 0.24400 | 0.28900 | 0.26100 |
| 9 | 0.28700 | 0.31000 | 0.28100 | 0.27600 | 0.28800 | 0.29100 | 0.28300 | 0.30100 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 0 | 0.01500 | 0.01200 | 0.01200 | 0.01400 | 0.01400 | 0.01100 | 0.01000 |
| 1 | 0.03300 | 0.04800 | 0.03700 | 0.03700 | 0.03600 | 0.04400 | 0.04900 |
| 2 | 0.11300 | 0.11800 | 0.11800 | 0.10400 | 0.10000 | 0.09900 | 0.11700 |
| 3 | 0.15700 | 0.14900 | 0.15300 | 0.15800 | 0.13800 | 0.15300 | 0.14400 |
| 4 | 0.17900 | 0.17700 | 0.17000 | 0.17400 | 0.18300 | 0.16600 | 0.17200 |
| 5 | 0.20100 | 0.19800 | 0.19900 | 0.18400 | 0.20100 | 0.20800 | 0.18100 |
| 6 | 0.21600 | 0.21300 | 0.21400 | 0.20500 | 0.21600 | 0.22300 | 0.22000 |
| 7 | 0.24600 | 0.23800 | 0.22800 | 0.22200 | 0.22800 | 0.24000 | 0.23700 |
| 8 | 0.27500 | 0.26700 | 0.25000 | 0.23200 | 0.24600 | 0.25700 | 0.23500 |
| 9 | 0.26200 | 0.28800 | 0.25200 | 0.25600 | 0.27200 | 0.27800 | 0.26200 |


| Weights at age in the stock ( Kg ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 |
| 0 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 |
| 1 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 |
| 2 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 |
| 3 | 0.18700 | 0.18700 | 0.18700 | 0.18700 | 0.18700 | 0.18700 | 0.18700 | 0.18700 |
| 4 | 0.22300 | 0.22300 | 0.22300 | 0.22300 | 0.22300 | 0.22300 | 0.22300 | 0.22300 |
| 5 | 0.23900 | 0.23900 | 0.23900 | 0.23900 | 0.23900 | 0.23900 | 0.23900 | 0.23900 |
| 6 | 0.27600 | 0.27600 | 0.27600 | 0.27600 | 0.27600 | 0.27600 | 0.27600 | 0.27600 |
| 7 | 0.29900 | 0.29900 | 0.29900 | 0.29900 | 0.29900 | 0.29900 | 0.29900 | 0.29900 |
| 8 | 0.30600 | 0.30600 | 0.30600 | 0.30600 | 0.30600 | 0.30600 | 0.30600 | 0.30600 |
| 9 | 0.31200 | 0.31200 | 0.31200 | 0.31200 | 0.31200 | 0.31200 | 0.31200 | 0.31200 |



| Weights at age in the stock ( Kg ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| 0 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01500 | 0.01700 |
| 1 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05000 | 0.05700 |
| 2 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15000 |
| 3 | 0.18700 | 0.18700 | 0.18700 | 0.18700 | 0.18700 | 0.18700 | 0.18700 | 0.19000 |
| 4 | 0.22300 | 0.22300 | 0.22300 | 0.22300 | 0.22300 | 0.22300 | 0.22300 | 0.23000 |
| 5 | 0.23900 | 0.23900 | 0.23900 | 0.23900 | 0.23900 | 0.23900 | 0.23900 | 0.24300 |
| 6 | 0.27600 | 0.27600 | 0.27600 | 0.27600 | 0.27600 | 0.27600 | 0.27600 | 0.28200 |
| 7 | 0.29900 | 0.29900 | 0.29900 | 0.29900 | 0.29900 | 0.29900 | 0.29900 | 0.31100 |
| 8 | 0.30600 | 0.30600 | 0.30600 | 0.30600 | 0.30600 | 0.30600 | 0.30600 | 0.33800 |
| 9 | 0.31200 | 0.31200 | 0.31200 | 0.31200 | 0.31200 | 0.31200 | 0.31200 | 0.34700 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Weights at age in the stock ( Kg ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 0 | 0.01600 | 0.01400 | 0.00900 | 0.00800 | 0.00800 | 0.01200 | 0.01100 | 0.01000 |
| 1 | 0.05600 | 0.06100 | 0.05000 | 0.04800 | 0.04400 | 0.05200 | 0.05900 | 0.06400 |
| 2 | 0.13800 | 0.13000 | 0.12200 | 0.12300 | 0.12200 | 0.12600 | 0.13900 | 0.13700 |
| 3 | 0.18700 | 0.18300 | 0.17000 | 0.16600 | 0.16500 | 0.17400 | 0.18400 | 0.19400 |
| 4 | 0.23200 | 0.23200 | 0.21200 | 0.20800 | 0.20500 | 0.21200 | 0.21200 | 0.21400 |
| 5 | 0.24700 | 0.25200 | 0.23000 | 0.22900 | 0.22800 | 0.24400 | 0.23900 | 0.23400 |
| 6 | 0.27500 | 0.27300 | 0.24200 | 0.24800 | 0.25200 | 0.27000 | 0.26500 | 0.25300 |
| 7 | 0.32100 | 0.31500 | 0.27500 | 0.25900 | 0.26100 | 0.28400 | 0.28000 | 0.27100 |
| 8 | 0.34100 | 0.33200 | 0.26800 | 0.26300 | 0.27700 | 0.29800 | 0.30000 | 0.29100 |
| 9 | 0.36500 | 0.39200 | 0.34300 | 0.32500 | 0.31500 | 0.33100 | 0.32800 | 0.31200 |


| Weights at age in the stock ( Kg ) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 0.00600 | 0.00700 | 0.00600 | 0.00600 | 0.00500 | 0.00600 | 0.00600 | 0.00600 |
| 1 | 0.06100 | 0.06000 | 0.05700 | 0.05400 | 0.04900 | 0.04700 | 0.05100 | 0.05100 |
| 2 | 0.13400 | 0.12700 | 0.13000 | 0.13000 | 0.12300 | 0.11600 | 0.11600 | 0.11600 |
| 3 | 0.18400 | 0.19200 | 0.18600 | 0.19900 | 0.18300 | 0.18700 | 0.17900 | 0.18400 |
| 4 | 0.21300 | 0.21400 | 0.21100 | 0.22800 | 0.23000 | 0.24100 | 0.22600 | 0.22100 |
| 5 | 0.23500 | 0.24000 | 0.22400 | 0.23400 | 0.23700 | 0.26400 | 0.25600 | 0.24800 |
| 6 | 0.26200 | 0.27500 | 0.26800 | 0.27400 | 0.25700 | 0.28400 | 0.27300 | 0.27900 |
| 7 | 0.27300 | 0.29100 | 0.29300 | 0.30100 | 0.28000 | 0.28700 | 0.27600 | 0.28600 |
| 8 | 0.30200 | 0.30900 | 0.31800 | 0.32400 | 0.30300 | 0.30100 | 0.27000 | 0.28100 |
| 9 | 0.32000 | 0.33800 | 0.34600 | 0.34400 | 0.33400 | 0.34200 | 0.31800 | 0.30300 |


| Weights at age in the stock ( Kg ) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 0 | 0.00600 | 0.00600 | 0.00700 | 0.00700 | 0.00600 | 0.00700 | 0.00600 |
| 1 | 0.05100 | 0.04700 | 0.04700 | 0.04200 | 0.04100 | 0.04100 | 0.04400 |
| 2 | 0.12200 | 0.12800 | 0.12300 | 0.11900 | 0.11800 | 0.12600 | 0.13100 |
| 3 | 0.17200 | 0.17200 | 0.17300 | 0.16500 | 0.16500 | 0.15500 | 0.16400 |
| 4 | 0.21000 | 0.20500 | 0.20200 | 0.20300 | 0.19800 | 0.19100 | 0.18400 |
| 5 | 0.23300 | 0.22800 | 0.22200 | 0.22300 | 0.22500 | 0.21600 | 0.20800 |
| 6 | 0.25500 | 0.24800 | 0.24200 | 0.24800 | 0.24800 | 0.24200 | 0.23600 |
| 7 | 0.27500 | 0.27000 | 0.26600 | 0.26800 | 0.26500 | 0.25200 | 0.24800 |
| 8 | 0.27400 | 0.28900 | 0.28500 | 0.28300 | 0.28100 | 0.26600 | 0.25900 |
| 9 | 0.28000 | 0.27500 | 0.28300 | 0.27500 | 0.29100 | 0.27700 | 0.28000 |


| Natural Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 |
| 0 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 9 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Natural Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 |
| 0 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 9 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |


| Natural Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| 0 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 9 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |


| Natural Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 0 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 9 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |


| Natural Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 9 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Natural Mortality (per year) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 0 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 9 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |


| Proportion of fish spawning |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 |
| 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 3 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |


| Proportion of fish spawning |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 |
| 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 0.8200 | 0.8200 | 0.8200 | 0.8200 |
| 3 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |


| Proportion of fish spawning |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.8200 | 0.8200 | 0.8200 | 0.8200 | 0.8200 | 0.8200 | 0.8200 | 0.8200 |
| 3 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Proportion of fish spawning |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.8200 | 0.7000 | 0.7500 | 0.8000 | 0.8500 | 0.8200 | 0.9100 | 0.8600 |
| 3 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 0.9300 | 0.9400 | 0.9700 | 0.9900 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |


| Proportion of fish spawning |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.5000 | 0.4700 | 0.7300 | 0.6700 | 0.6100 | 0.6400 | 0.6400 | 0.6900 |
| 3 | 0.9900 | 0.6100 | 0.9300 | 0.9500 | 0.9800 | 0.9400 | 0.8900 | 0.9100 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Proportion of fish spawning

| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.6700 | 0.7700 | 0.8700 | 0.4300 | 0.7000 | 0.7600 | 0.6600 |
| 3 | 0.9600 | 0.9200 | 0.9700 | 0.9300 | 0.6500 | 0.9600 | 0.8800 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 0.9600 | 0.9800 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

INDICES OF SPAWNING BIOMASS

| MLAI |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| 1 | 13.10 | 7.94 | 2.73 | 2.46 | 6.11 | 7.37 | 14.43 | 9.86 |
| MLAI |  |  |  |  |  |  |  |  |
|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| 1 | 14.54 | 21.03 | 26.97 | 48.82 | 74.35 | 38.63 | 67.99 | 134.91 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| MLAI |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 131.57 | 170.59 | 89.67 | 41.88 | 29.85 | 20.65 | 22.74 | 44.73 |
| MLAI |  |  |  |  |  |  |  |  |
|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 1 | 57.25 | 74.39 | 61.91 | 41.09 | 130.31 | 110.96 | 271.65 | 321.99 |
| MLAI |  |  |  |  |  |  |  |  |
|  | 2005 | 2006 |  |  |  |  |  |  |
| 1 | 192.71 | 118.40 |  |  |  |  |  |  |

## AGE-STRUCTURED INDICES

| Acoustic survey 1-9+ wr |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 999990. | 999990. | 999990. | 999990. | 999990. | 999990. | 999990. | 999990. |
| 2 | 4090. | 3306. | 2634. | 3734. | 2984. | 3185. | 3849. | 4497. |
| 3 | 3903. | 3521. | 1700. | 1378. | 1637. | 839. | 2041 | 2824. |
| 4 | 1633. | 3414. | 1959. | 1147. | 902. | 399. | 672. | 1087. |
| 5 | 492. | 1366. | 1849. | 1134. | 741. | 381. | 299. | 311. |
| 6 | 283. | 392. | 644. | 1246. | 777. | 321. | 203. | 99. |
| 7 | 120. | 210. | 228. | 395. | 551. | 326. | 138. | 83. |
| 8 | 44. | 133. | 94. | 114. | 180. | 219. | 119. | 133. |
| 9 | 22. | 43. | 51. | 104. | 116. | 131. | 93. | 206. |


| AGE | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 9361. | 4449. | 5087. | 24736. | 6837. | 23055. | 9829. | 5184. |
| 2 | 5960. | 5747. | 3078. | 2923. | 12290. | 4875. | 18949. | 3416. |
| 3 | 2935. | 2520. | 4725. | 2156. | 3083. | 8220. | 3081. | 9192. |
| 4 | 1441. | 1625. | 1116. | 3140. | 1462. | 1390. | 4189. | 2167. |
| 5 | 601. | 982. | 506. | 1007. | 1676. | 795. | 675. | 2591. |
| 6 | 215. | 445. | 314. | 483. | 450. | 1031. | 495. | 317. |
| 7 | 46. | 170. | 139. | 266. | 170. | 244. | 568. | 328. |
| 8 | 78. | 45. | 54. | 120. | 98. | 121. | 146. | 342. |
| 9 | 159. | 121. | 87. | 97. | 59. | 149. | 178. | 186. |

$x 10 \wedge 3$
Acoustic survey 1-9+ wr

| AGE | 2005 | 2006 |
| :---: | :---: | :---: |
| 1 | 3114. | 6823. |
| 2 | 2055. | 3772. |
| 3 | 3649. | 1997. |
| 4 | 5790. | 2097. |
| 5 | 1213. | 4175. |
| 6 | 1175. | 618. |
| 7 | 140. | 562. |
| 8 | 126. | 84. |
| 9 | 107. | 70. |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| IBTS1: 1-5+ wr |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 1 | 1515.6 | 2097.3 | 2662.8 | 3693.0 | 4394.2 | 2331.6 | 1061.6 | 1286.7 |
| 2 | 161.5 | 721.6 | 782.1 | 917.5 | 4163.4 | 875.3 | 462.1 | 693.0 |
| 3 | 61.4 | 282.0 | 276.0 | 116.3 | 791.5 | 338.5 | 279.8 | 258.6 |
| 4 | 26.9 | 42.1 | 79.0 | 59.4 | 58.0 | 89.4 | 269.1 | 221.5 |
| 5 | 10.2 | 27.9 | 28.1 | 48.8 | 25.1 | 8.5 | 71.3 | 146.1 |
| IBTS1: 1-5+ wr |  |  |  |  |  |  |  |  |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 1 | 1268.1 | 2794.0 | 1752.1 | 1345.8 | 1890.9 | 4404.6 | 2275.8 | 752.9 |
| 2 | 436.6 | 787.4 | 1167.2 | 1392.9 | 197.5 | 506.5 | 791.6 | 450.6 |
| 3 | 193.1 | 222.6 | 213.1 | 278.5 | 32.9 | 162.7 | 95.7 | 501.3 |
| 4 | 54.8 | 45.0 | 69.0 | 36.7 | 10.2 | 30.5 | 20.8 | 98.2 |
| 5 | 92.3 | 65.5 | 42.5 | 6.6 | 8.1 | 19.9 | 17.8 | 35.6 |
| IBTS1: 1-5+ wr |  |  |  |  |  |  |  |  |
| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 1 | 3725.1 | 2499.4 | 4064.8 | 2765.1 | 979.0 | 1001.6 | 922.0 | 1336.3 |
| 2 | 199.4 | 1129.3 | 658.2 | 1556.1 | 451.0 | 214.2 | 1464.3 | 40.7 |
| 3 | 154.7 | 317.1 | 338.2 | 611.9 | 777.3 | 356.0 | 330.0 | 18.2 |
| 4 | 58.8 | 93.9 | 25.0 | 360.0 | 112.4 | 388.9 | 251.7 | 8.4 |
| 5 | 9.0 | 68.3 | 19.9 | 53.2 | 171.2 | 131.5 | 338.8 | 40.9 |
| MIK 0-wr |  |  |  |  |  |  |  |  |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 200.70 | 190.10 | 101.70 | 127.00 | 106.50 | 148.10 | 53.10 | 244.00 |
| MIK 0-wr |  |  |  |  |  |  |  |  |
| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 0 | 137.10 | 214.80 | 161.80 | 54.40 | 47.30 | 61.30 | 83.10 | 37.20 |


| Fishing Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 |
| 0 | 0.0257 | 0.0186 | 0.0049 | 0.0148 | 0.0126 | 0.0071 | 0.0215 | 0.0256 |
| 1 | 0.2557 | 0.1294 | 0.0897 | 0.1241 | 0.3084 | 0.2461 | 0.1852 | 0.2980 |
| 2 | 0.4364 | 0.6166 | 0.2502 | 0.2975 | 0.3889 | 0.7753 | 0.5921 | 0.4222 |
| 3 | 0.3285 | 0.3527 | 0.6260 | 0.2755 | 0.4123 | 0.7387 | 0.7082 | 0.8046 |
| 4 | 0.3375 | 0.4088 | 0.4220 | 0.2264 | 0.3702 | 0.7766 | 0.5716 | 0.9244 |
| 5 | 0.2666 | 0.4025 | 0.5348 | 0.1507 | 0.3068 | 0.6598 | 0.8343 | 0.8271 |
| 6 | 0.3130 | 0.3821 | 0.8179 | 0.1820 | 0.2379 | 0.5172 | 0.3903 | 1.0090 |
| 7 | 0.6088 | 0.2498 | 0.6379 | 0.2850 | 0.2824 | 0.4569 | 0.3852 | 1.5312 |
| 8 | 0.5634 | 0.5322 | 0.5790 | 0.3372 | 0.5568 | 0.8634 | 0.7312 | 1.0509 |
| 9 | 0.5634 | 0.5322 | 0.5790 | 0.3372 | 0.5568 | 0.8634 | 0.7312 | 1.0509 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Fishing Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 |
| 0 | 0.0348 | 0.0082 | 0.0351 | 0.0340 | 0.0583 | 0.0462 | 0.0749 | 0.1570 |
| 1 | 0.3002 | 0.3291 | 0.2681 | 0.6021 | 0.5781 | 0.6739 | 0.4514 | 0.6880 |
| 2 | 1.3272 | 0.7844 | 0.9728 | 0.8826 | 0.8121 | 1.0219 | 1.0287 | 1.3100 |
| 3 | 1.8720 | 0.9124 | 1.2669 | 1.2147 | 0.8014 | 1.3335 | 0.9725 | 1.5045 |
| 4 | 1.0715 | 0.8741 | 1.3303 | 1.2263 | 0.7996 | 0.9877 | 0.9932 | 1.3707 |
| 5 | 1.2340 | 1.0541 | 0.8755 | 1.0843 | 0.5494 | 0.9514 | 1.1856 | 1.8787 |
| 6 | 1.1729 | 1.9008 | 1.0800 | 2.6145 | 0.5173 | 1.3770 | 1.0784 | 1.2742 |
| 7 | 1.5948 | 1.2928 | 4.1124 | 2.7132 | 0.0981 | 0.8048 | 0.7714 | 2.0312 |
| 8 | 1.6467 | 1.3070 | 1.7058 | 1.9039 | 1.0387 | 1.5539 | 1.3299 | 2.0080 |
| 9 | 1.6467 | 1.3070 | 1.7058 | 1.9039 | 1.0387 | 1.5539 | 1.3299 | 2.0080 |


| Fishing Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| 0 | 0.1465 | 0.0975 | 0.0455 | 0.0837 | 0.1257 | 0.4818 | 0.3343 | 0.3996 |
| 1 | 0.2486 | 0.2966 | 0.2000 | 0.1665 | 0.1132 | 0.2853 | 0.2250 | 0.2516 |
| 2 | 1.3390 | 0.2245 | 0.0242 | 0.0947 | 0.3634 | 0.3241 | 0.2605 | 0.3020 |
| 3 | 1.4294 | 1.4112 | 0.0424 | 0.0663 | 0.4191 | 0.2751 | 0.5083 | 0.3243 |
| 4 | 1.7396 | 0.4259 | 0.1041 | 0.0935 | 0.2965 | 0.3034 | 0.2468 | 0.4365 |
| 5 | 1.5898 | 1.2081 | 0.0165 | 0.0523 | 0.2645 | 0.4114 | 0.1543 | 0.2751 |
| 6 | 1.0714 | 0.7257 | 0.0777 | 0.0123 | 0.0672 | 0.4300 | 0.1444 | 0.3446 |
| 7 | 1.4994 | 0.7402 | 0.0595 | 0.4405 | 0.1008 | 0.9674 | 0.2286 | 0.3895 |
| 8 | 1.6429 | 0.9525 | 0.1771 | 0.2270 | 0.3653 | 0.6135 | 0.4286 | 0.5099 |
| 9 | 1.6429 | 0.9525 | 0.1771 | 0.2270 | 0.3653 | 0.6135 | 0.4286 | 0.5099 |


| Fishing Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 0 | 0.2263 | 0.0852 | 0.0619 | 0.1614 | 0.1247 | 0.1303 | 0.0589 | 0.1179 |
| 1 | 0.2051 | 0.3827 | 0.3157 | 0.3723 | 0.5800 | 0.4308 | 0.4528 | 0.3082 |
| 2 | 0.3144 | 0.4043 | 0.4592 | 0.4061 | 0.3556 | 0.3983 | 0.3769 | 0.5743 |
| 3 | 0.4295 | 0.6708 | 0.5225 | 0.5053 | 0.4006 | 0.4100 | 0.3695 | 0.4546 |
| 4 | 0.5368 | 0.7371 | 0.5814 | 0.5890 | 0.5814 | 0.5553 | 0.4674 | 0.4576 |
| 5 | 0.6273 | 0.6630 | 0.5532 | 0.6156 | 0.6641 | 0.6555 | 0.4994 | 0.4833 |
| 6 | 0.3590 | 0.7298 | 0.7302 | 0.6341 | 0.6728 | 0.7005 | 0.4911 | 0.4769 |
| 7 | 0.6955 | 0.5551 | 0.8164 | 0.6088 | 0.6882 | 0.7031 | 0.6797 | 0.4210 |
| 8 | 0.6080 | 0.8591 | 0.8008 | 0.7866 | 0.9017 | 0.8230 | 0.7593 | 0.7055 |
| 9 | 0.6080 | 0.8591 | 0.8008 | 0.7866 | 0.9017 | 0.8230 | 0.7593 | 0.7055 |


| Fishing Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 0.2968 | 0.3762 | 0.2285 | 0.3231 | 0.0754 | 0.0248 | 0.0152 | 0.0364 |
| 1 | 0.3873 | 0.4222 | 0.2462 | 0.2969 | 0.2551 | 0.0453 | 0.1663 | 0.0499 |
| 2 | 0.5726 | 0.6688 | 0.6838 | 0.6004 | 0.3140 | 0.2888 | 0.2661 | 0.2595 |
| 3 | 0.4983 | 0.6407 | 0.7167 | 0.8672 | 0.4909 | 0.4016 | 0.4111 | 0.4159 |
| 4 | 0.5727 | 0.7334 | 0.9111 | 0.8682 | 0.4186 | 0.5129 | 0.4409 | 0.3923 |
| 5 | 0.5463 | 0.7112 | 0.5574 | 0.8229 | 0.4794 | 0.4530 | 0.6099 | 0.3579 |
| 6 | 0.7208 | 0.6991 | 0.6693 | 0.5397 | 0.3139 | 0.4629 | 0.7231 | 0.4565 |
| 7 | 0.6940 | 0.8759 | 0.4760 | 0.6445 | 0.1431 | 0.2410 | 0.3797 | 0.4619 |
| 8 | 0.8558 | 1.0015 | 0.8593 | 0.9310 | 0.5374 | 0.4211 | 0.5544 | 0.4168 |
| 9 | 0.8558 | 1.0015 | 0.8593 | 0.9310 | 0.5374 | 0.4211 | 0.5544 | 0.4168 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Fishing Mortality (per year) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 0 | 0.0446 | 0.0322 | 0.0434 | 0.0441 | 0.0501 | 0.0635 | 0.0604 |
| 1 | 0.0781 | 0.0713 | 0.0655 | 0.0666 | 0.0756 | 0.0959 | 0.0912 |
| 2 | 0.2336 | 0.1240 | 0.1374 | 0.1398 | 0.1587 | 0.2012 | 0.1914 |
| 3 | 0.3378 | 0.3068 | 0.2017 | 0.2053 | 0.2330 | 0.2955 | 0.2810 |
| 4 | 0.4583 | 0.3317 | 0.2894 | 0.2946 | 0.3343 | 0.4240 | 0.4032 |
| 5 | 0.4863 | 0.3853 | 0.3184 | 0.3241 | 0.3678 | 0.4664 | 0.4436 |
| 6 | 0.3414 | 0.3497 | 0.3219 | 0.3277 | 0.3719 | 0.4716 | 0.4485 |
| 7 | 0.3998 | 0.2341 | 0.3139 | 0.3195 | 0.3626 | 0.4598 | 0.4373 |
| 8 | 0.4140 | 0.3184 | 0.2894 | 0.2946 | 0.3343 | 0.4240 | 0.4032 |
| 9 | 0.4140 | 0.3184 | 0.2894 | 0.2946 | 0.3343 | 0.4240 | 0.4032 |


| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 |
| 0 | 12.09 | 108.86 | 46.28 | 47.66 | 62.79 | 34.90 | 27.86 | 40.26 |
| 1 | 16.43 | 4.33 | 39.31 | 16.94 | 17.27 | 22.81 | 12.75 | 10.03 |
| 2 | 3.70 | 4.68 | 1.40 | 13.22 | 5.51 | 4.67 | 6.56 | 3.90 |
| 3 | 7.71 | 1.77 | 1.87 | 0.81 | 7.27 | 2.76 | 1.59 | 2.69 |
| 4 | 0.61 | 4.54 | 1.02 | 0.82 | 0.50 | 3.94 | 1.08 | 0.64 |
| 5 | 0.75 | 0.39 | 2.73 | 0.60 | 0.59 | 0.31 | 1.64 | 0.55 |
| 6 | 0.44 | 0.52 | 0.24 | 1.45 | 0.47 | 0.39 | 0.15 | 0.64 |
| 7 | 0.29 | 0.29 | 0.32 | 0.09 | 1.09 | 0.34 | 0.21 | 0.09 |
| 8 | 0.31 | 0.14 | 0.21 | 0.15 | 0.06 | 0.75 | 0.19 | 0.13 |
| 9 | 0.34 | 0.22 | 0.21 | 0.19 | 0.14 | 0.14 | 0.48 | 0.28 |


| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 |
| 0 | 38.70 | 21.58 | 41.08 | 32.31 | 20.86 | 10.11 | 21.70 | 2.84 |
| 1 | 14.43 | 13.75 | 7.87 | 14.59 | 11.49 | 7.24 | 3.55 | 7.41 |
| 2 | 2.74 | 3.93 | 3.64 | 2.22 | 2.94 | 2.37 | 1.36 | 0.83 |
| 3 | 1.89 | 0.54 | 1.33 | 1.02 | 0.68 | 0.97 | 0.63 | 0.36 |
| 4 | 0.98 | 0.24 | 0.18 | 0.31 | 0.25 | 0.25 | 0.21 | 0.20 |
| 5 | 0.23 | 0.31 | 0.09 | 0.04 | 0.08 | 0.10 | 0.08 | 0.07 |
| 6 | 0.22 | 0.06 | 0.10 | 0.03 | 0.01 | 0.04 | 0.04 | 0.02 |
| 7 | 0.21 | 0.06 | 0.01 | 0.03 | 0.00 | 0.01 | 0.01 | 0.01 |
| 8 | 0.02 | 0.04 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 9 | 0.12 | 0.04 | 0.02 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 |


| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| 0 | 2.73 | 4.34 | 4.61 | 10.61 | 16.74 | 37.88 | 64.78 | 61.83 |
| 1 | 0.89 | 0.87 | 1.45 | 1.62 | 3.59 | 5.43 | 8.61 | 17.06 |
| 2 | 1.37 | 0.26 | 0.24 | 0.44 | 0.50 | 1.18 | 1.50 | 2.53 |
| 3 | 0.17 | 0.27 | 0.15 | 0.17 | 0.29 | 0.26 | 0.63 | 0.86 |
| 4 | 0.07 | 0.03 | 0.05 | 0.12 | 0.13 | 0.16 | 0.16 | 0.31 |
| 5 | 0.04 | 0.01 | 0.02 | 0.04 | 0.10 | 0.09 | 0.11 | 0.11 |
| 6 | 0.01 | 0.01 | 0.00 | 0.02 | 0.04 | 0.07 | 0.05 | 0.08 |
| 7 | 0.01 | 0.00 | 0.00 | 0.00 | 0.02 | 0.03 | 0.04 | 0.04 |
| 8 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.03 |
| 9 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.03 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 0 | 53.48 | 80.96 | 97.63 | 86.22 | 42.29 | 39.17 | 35.87 | 33.63 |
| 1 | 15.25 | 15.69 | 27.35 | 33.76 | 26.99 | 13.73 | 12.65 | 12.44 |
| 2 | 4.88 | 4.57 | 3.94 | 7.34 | 8.56 | 5.56 | 3.28 | 2.96 |
| 3 | 1.39 | 2.64 | 2.26 | 1.84 | 3.62 | 4.44 | 2.77 | 1.67 |
| 4 | 0.51 | 0.74 | 1.11 | 1.10 | 0.91 | 1.99 | 2.41 | 1.56 |
| 5 | 0.18 | 0.27 | 0.32 | 0.56 | 0.55 | 0.46 | 1.03 | 1.37 |
| 6 | 0.08 | 0.09 | 0.13 | 0.17 | 0.27 | 0.26 | 0.22 | 0.57 |
| 7 | 0.05 | 0.05 | 0.04 | 0.05 | 0.08 | 0.13 | 0.12 | 0.12 |
| 8 | 0.03 | 0.02 | 0.03 | 0.02 | 0.03 | 0.04 | 0.06 | 0.05 |
| 9 | 0.04 | 0.03 | 0.03 | 0.02 | 0.01 | 0.02 | 0.02 | 0.03 |


| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 62.13 | 50.24 | 34.19 | 41.51 | 50.07 | 27.75 | 27.13 | 69.07 |
| 1 | 11.00 | 16.99 | 12.69 | 10.01 | 11.06 | 17.08 | 9.96 | 9.83 |
| 2 | 3.36 | 2.75 | 4.10 | 3.65 | 2.74 | 3.15 | 6.01 | 3.10 |
| 3 | 1.23 | 1.41 | 1.04 | 1.53 | 1.48 | 1.48 | 1.75 | 3.41 |
| 4 | 0.87 | 0.61 | 0.61 | 0.42 | 0.53 | 0.74 | 0.81 | 0.95 |
| 5 | 0.90 | 0.44 | 0.27 | 0.22 | 0.16 | 0.31 | 0.40 | 0.47 |
| 6 | 0.76 | 0.47 | 0.20 | 0.14 | 0.09 | 0.09 | 0.18 | 0.20 |
| 7 | 0.32 | 0.34 | 0.21 | 0.09 | 0.07 | 0.06 | 0.05 | 0.08 |
| 8 | 0.07 | 0.14 | 0.13 | 0.12 | 0.04 | 0.06 | 0.04 | 0.03 |
| 9 | 0.04 | 0.07 | 0.08 | 0.05 | 0.01 | 0.02 | 0.01 | 0.01 |


| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 0 | 39.91 | 91.32 | 31.39 | 18.56 | 22.56 | 17.51 | 27.78 | 11.92 |
| 1 | 24.50 | 14.04 | 32.53 | 11.06 | 6.53 | 7.89 | 6.04 | 9.62 |
| 2 | 3.44 | 8.34 | 4.81 | 11.21 | 3.81 | 2.23 | 2.64 | 2.03 |
| 3 | 1.77 | 2.02 | 5.45 | 3.11 | 7.22 | 2.41 | 1.35 | 1.61 |
| 4 | 1.84 | 1.04 | 1.22 | 3.65 | 2.07 | 4.68 | 1.47 | 0.83 |
| 5 | 0.58 | 1.05 | 0.67 | 0.82 | 2.46 | 1.34 | 2.77 | 0.89 |
| 6 | 0.30 | 0.32 | 0.65 | 0.44 | 0.54 | 1.54 | 0.76 | 1.61 |
| 7 | 0.11 | 0.19 | 0.21 | 0.43 | 0.29 | 0.34 | 0.87 | 0.44 |
| 8 | 0.05 | 0.07 | 0.14 | 0.14 | 0.28 | 0.18 | 0.19 | 0.51 |
| 9 | 0.01 | 0.01 | 0.04 | 0.04 | 0.12 | 0.18 | 0.01 | 0.12 |

$x 10 \wedge 9$

| AGE | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 1 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 2 | 3.6700 | 3.6700 | 3.6700 | 3.6700 | 3.6700 |
| 3 | 2.8700 | 2.8700 | 2.8700 | 2.8700 | 2.8700 |
| 4 | 2.2300 | 2.2300 | 2.2300 | 2.2300 | 2.2300 |
| 5 | 1.7400 | 1.7400 | 1.7400 | 1.7400 | 1.7400 |
| 6 | 1.3700 | 1.3700 | 1.3700 | 1.3700 | 1.3700 |
| 7 | 1.0400 | 1.0400 | 1.0400 | 1.0400 | 1.0400 |
| 8 | 0.9400 | 0.9400 | 0.9400 | 0.9400 | 0.9400 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| MLAI |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| 1 | 16.75 | 11.02 | 5.04 | 4.78 | 2.73 | 3.89 | 6.90 | 8.69 |


|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 13.75 | 20.60 | 34.13 | 57.21 | 59.16 | 57.23 | 79.04 | 109.22 |


|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 115.03 | 108.25 | 87.05 | 59.58 | 37.84 | 41.26 | 36.68 | 36.44 |


| MLAI |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 1 | 45.18 | 61.89 | 73.44 | 73.52 | 119.69 | 150.58 | 163.76 | 169.95 |


| MLAI |  |  |
| :---: | :---: | :---: |
|  | 2005 | 2006 |
| 1 | 151.64 | 110.38 |

## Predicted Age-Structured Index Values

| Acoustic survey 1-9+ wr Predicted |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 999990. | 999990. | 999990. | 999990. | 999990. | 999990. | 999990. | 999990. |
| 2 | 5892. | 3521. | 2847. | 3239. | 2508. | 3712. | 3460. | 3038. |
| 3 | 5749. | 3659. | 2107. | 1521. | 1602. | 1139. | 1542. | 1835. |
| 4 | 2549. | 3250. | 2118. | 1102. | 714. | 640. | 450. | 729. |
| 5 | 579. | 1414. | 1893. | 1197. | 540. | 354. | 253. | 219. |
| 6 | 310. | 293. | 773. | 912. | 567. | 241. | 182. | 131. |
| 7 | 141. | 130. | 156. | 357. | 342. | 267. | 105. | 111. |
| 8 | 42. | 68. | 66. | 81. | 152. | 145. | 130. | 59. |
| 9 | 52. | 77. | 91. | 132. | 198. | 263. | 151. | 37. |


| Acoustic survey 1-9+ wr Predicted |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 1 | 11351. | 6191. | 6516. | 15990. | 9199. | 21380. | 7262. | 4269. |
| 2 | 3547 . | 6845. | 3548. | 3992. | 10273. | 5885. | 13695. | 4601. |
| 3 | 1925. | 2262. | 4398. | 2387. | 2763. | 7915. | 4498. | 10298. |
| 4 | 976. | 1108. | 1332. | 2492. | 1502. | 1805. | 5405. | 3000. |
| 5 | 441. | 519. | 700. | 801. | 1538. | 1018. | 1243. | 3625. |
| 6 | 122. | 215. | 273. | 439. | 473. | 964. | 656. | 779. |
| 7 | 84. | 67. | 101. | 150. | 278. | 285. | 587. | 389. |
| 8 | 83. | 56. | 46. | 66. | 106. | 215. | 212. | 426. |
| 9 | 69. | 34. | 25. | 30. | 44. | 155. | 158. | 489. |


|  | Acoustic survey |  |
| :---: | :---: | :---: |
| AGE | 2005 | 2006 |
| 1 | 5102. | 3916. |
| 2 | 2632. | 3133. |
| 3 | 3315. | 1875. |
| 4 | 6457. | 2044. |
| 5 | 1872. | 3919. |
| 6 | 2109. | 1056. |
| 7 | 430. | 1126. |
| 8 | 264. | 282. |
| 9 | 704. | 55. |


| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2140.2 | 2153.2 | 3785.0 | 4638.9 | 3613.9 | 1873.3 | 1720.8 | 1723.6 |
| 2 | 701.8 | 649.9 | 556.0 | 1043.3 | 1224.5 | 791.2 | 468.6 | 411.9 |
| 3 | 139.9 | 258.7 | 225.6 | 184.3 | 367.2 | 449.9 | 281.4 | 168.0 |
| 4 | 29.5 | 41.8 | 63.9 | 63.3 | 52.6 | 115.2 | 141.5 | 91.8 |
| 5 | 11.8 | 14.0 | 16.5 | 25.0 | 28.9 | 27.4 | 45.0 | 66.9 |


| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1508.2 | 2320.0 | 1770.9 | 1388.4 | 1541.5 | 2445.1 | 1403.9 | 1406.4 |
| 2 | 468.3 | 377.8 | 562.7 | 506.3 | 393.5 | 454.6 | 868.9 | 449.1 |
| 3 | 123.6 | 138.3 | 101.6 | 146.5 | 148.6 | 150.1 | 177.0 | 345.0 |
| 4 | 50.2 | 34.8 | 33.6 | 23.2 | 31.1 | 43.3 | 47.7 | 56.2 |
| 5 | 64.2 | 44.1 | 27.3 | 18.7 | 11.8 | 16.9 | 21.1 | 25.0 |


| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 3492.8 | 2003.6 | 4645.1 | 1578.6 | 931.5 | 1122.9 | 860.2 | 1369.1 |
| 2 | 499.8 | 1227.6 | 707.3 | 1647.6 | 558.0 | 325.0 | 385.3 | 296.4 |
| 3 | 181.1 | 207.0 | 566.9 | 322.7 | 747.4 | 247.1 | 138.9 | 166.1 |
| 4 | 108.1 | 61.7 | 72.9 | 218.6 | 123.4 | 276.0 | 86.6 | 49.3 |
| 5 | 33.0 | 52.5 | 54.5 | 59.7 | 117.3 | 112.5 | 145.3 | 112.6 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| MIK 0-wr Predicted |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 188.20 | 150.65 | 104.46 | 125.33 | 155.90 | 86.95 | 85.12 | 216.13 |


| MIK 0-wr Predicted |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 0 | 124.77 | 285.91 | 98.13 | 58.01 | 70.47 | 54.59 | 86.66 | 37.20 |


| AGE | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.0763 | 0.0455 | 0.0115 | 0.0653 | 0.0340 | 0.0092 | 0.0375 | 0.0277 |
| 1 | 0.7578 | 0.3165 | 0.2125 | 0.5479 | 0.8331 | 0.3169 | 0.3241 | 0.3224 |
| 2 | 1.2931 | 1.5084 | 0.5927 | 1.3138 | 1.0505 | 0.9984 | 1.0358 | 0.4567 |
| 3 | 0.9733 | 0.8628 | 1.4832 | 1.2165 | 1.1137 | 0.9513 | 1.2390 | 0.8703 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.7899 | 0.9848 | 1.2672 | 0.6655 | 0.8285 | 0.8497 | 1.4596 | 0.8947 |
| 6 | 0.9276 | 0.9347 | 1.9379 | 0.8039 | 0.6425 | 0.6660 | 0.6827 | 1.0915 |
| 7 | 1.8042 | 0.6111 | 1.5114 | 1.2587 | 0.7626 | 0.5884 | 0.6738 | 1.6563 |
| 8 | 1.6694 | 1.3019 | 1.3719 | 1.4890 | 1.5039 | 1.1118 | 1.2791 | 1.1368 |
| 9 | 1.6694 | 1.3019 | 1.3719 | 1.4890 | 1.5039 | 1.1118 | 1.2791 | 1.1368 |


| Fitted Selection Pattern |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 |
| 0 | 0.0325 | 0.0094 | 0.0264 | 0.0277 | 0.0729 | 0.0467 | 0.0754 | 0.1145 |
| 1 | 0.2802 | 0.3765 | 0.2015 | 0.4910 | 0.7230 | 0.6823 | 0.4545 | 0.5019 |
| 2 | 1.2387 | 0.8973 | 0.7313 | 0.7197 | 1.0156 | 1.0347 | 1.0357 | 0.9557 |
| 3 | 1.7472 | 1.0438 | 0.9524 | 0.9906 | 1.0022 | 1.3501 | 0.9791 | 1.0976 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.1517 | 1.2059 | 0.6582 | 0.8842 | 0.6870 | 0.9633 | 1.1938 | 1.3706 |
| 6 | 1.0946 | 2.1745 | 0.8119 | 2.1321 | 0.6470 | 1.3942 | 1.0858 | 0.9296 |
| 7 | 1.4884 | 1.4790 | 3.0915 | 2.2126 | 0.1227 | 0.8149 | 0.7767 | 1.4819 |
| 8 | 1.5369 | 1.4952 | 1.2823 | 1.5527 | 1.2990 | 1.5733 | 1.3390 | 1.4650 |
| 9 | 1.5369 | 1.4952 | 1.2823 | 1.5527 | 1.2990 | 1.5733 | 1.3390 | 1.4650 |


| AGE | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.0842 | 0.2289 | 0.4368 | 0.8949 | 0.4240 | 1.5881 | 1.3544 | 0.9155 |
| 1 | 0.1429 | 0.6963 | 1.9209 | 1.7810 | 0.3818 | 0.9404 | 0.9114 | 0.5765 |
| 2 | 0.7697 | 0.5272 | 0.2320 | 1.0125 | 1.2258 | 1.0681 | 1.0552 | 0.6920 |
| 3 | 0.8217 | 3.3132 | 0.4072 | 0.7095 | 1.4135 | 0.9067 | 2.0594 | 0.7431 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.9139 | 2.8363 | 0.1584 | 0.5594 | 0.8920 | 1.3558 | 0.6252 | 0.6302 |
| 6 | 0.6159 | 1.7037 | 0.7466 | 0.1319 | 0.2266 | 1.4173 | 0.5852 | 0.7896 |
| 7 | 0.8619 | 1.7379 | 0.5717 | 4.7115 | 0.3400 | 3.1885 | 0.9263 | 0.8923 |
| 8 | 0.9445 | 2.2363 | 1.7010 | 2.4279 | 1.2323 | 2.0221 | 1.7364 | 1.1682 |
| 9 | 0.9445 | 2.2363 | 1.7010 | 2.4279 | 1.2323 | 2.0221 | 1.7364 | 1.1682 |

Table 2.6.4 (Cont) North Sea herring. Final model fit ICA output. Note age=ringer

| Fitted Selection Pattern |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 0 | 0.4215 | 0.1156 | 0.1065 | 0.2739 | 0.2144 | 0.2347 | 0.1260 | 0.2576 |
| 1 | 0.3822 | 0.5192 | 0.5429 | 0.6320 | 0.9975 | 0.7757 | 0.9687 | 0.6734 |
| 2 | 0.5856 | 0.5485 | 0.7899 | 0.6894 | 0.6116 | 0.7172 | 0.8065 | 1.2549 |
| 3 | 0.8000 | 0.9101 | 0.8986 | 0.8578 | 0.6889 | 0.7384 | 0.7906 | 0.9934 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.1686 | 0.8994 | 0.9515 | 1.0450 | 1.1421 | 1.1803 | 1.0685 | 1.0561 |
| 6 | 0.6688 | 0.9901 | 1.2559 | 1.0765 | 1.1572 | 1.2614 | 1.0508 | 1.0422 |
| 7 | 1.2956 | 0.7531 | 1.4042 | 1.0336 | 1.1835 | 1.2662 | 1.4544 | 0.9200 |
| 8 | 1.1327 | 1.1655 | 1.3774 | 1.3354 | 1.5509 | 1.4820 | 1.6246 | 1.5416 |
| 9 | 1.1327 | 1.1655 | 1.3774 | 1.3354 | 1.5509 | 1.4820 | 1.6246 | 1.5416 |


| Fitted Selection Pattern |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 0.5182 | 0.5130 | 0.2508 | 0.3722 | 0.1800 | 0.0484 | 0.0345 | 0.0928 |
| 1 | 0.6762 | 0.5757 | 0.2702 | 0.3419 | 0.6094 | 0.0883 | 0.3772 | 0.1271 |
| 2 | 0.9999 | 0.9119 | 0.7505 | 0.6915 | 0.7502 | 0.5632 | 0.6034 | 0.6614 |
| 3 | 0.8701 | 0.8736 | 0.7866 | 0.9989 | 1.1728 | 0.7831 | 0.9323 | 1.0601 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.9539 | 0.9698 | 0.6118 | 0.9478 | 1.1453 | 0.8834 | 1.3832 | 0.9123 |
| 6 | 1.2586 | 0.9532 | 0.7346 | 0.6217 | 0.7500 | 0.9026 | 1.6398 | 1.1636 |
| 7 | 1.2119 | 1.1943 | 0.5224 | 0.7424 | 0.3419 | 0.4698 | 0.8611 | 1.1774 |
| 8 | 1.4943 | 1.3656 | 0.9432 | 1.0723 | 1.2840 | 0.8211 | 1.2574 | 1.0624 |
| 9 | 1.4943 | 1.3656 | 0.9432 | 1.0723 | 1.2840 | 0.8211 | 1.2574 | 1.0624 |

Fitted Selection Pattern

| AGE | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.0974 | 0.0970 | 0.1498 | 0.1498 | 0.1498 | 0.1498 | 0.1498 |
| 1 | 0.1705 | 0.2150 | 0.2262 | 0.2262 | 0.2262 | 0.2262 | 0.2262 |
| 2 | 0.5098 | 0.3739 | 0.4747 | 0.4747 | 0.4747 | 0.4747 | 0.4747 |
| 3 | 0.7371 | 0.9250 | 0.6970 | 0.6970 | 0.6970 | 0.6970 | 0.6970 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0611 | 1.1617 | 1.1002 | 1.1002 | 1.1002 | 1.1002 | 1.1002 |
| 6 | 0.7450 | 1.0544 | 1.1124 | 1.1124 | 1.1124 | 1.1124 | 1.1124 |
| 7 | 0.8725 | 0.7059 | 1.0846 | 1.0846 | 1.0846 | 1.0846 | 1.0846 |
| 8 | 0.9034 | 0.9600 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 0.9034 | 0.9600 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Table 2.6.5 North Sea herring. STOCK SUMMARY


Table 2.6.6 North Sea herring. Model fit parameters, residuals and diagnostics.

## PARAMETER ESTIMATES

| ${ }^{3}$ Parm. ${ }^{3}$ |  | Maximum |  | ${ }^{3}$ | 3 |  | 3 | Mean of |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ${ }_{3}$ No. ${ }^{3}$ |  | Likelh. |  | Lower | Upper | -s.e. | +s.e. | Param. |
| ${ }^{3} \quad{ }^{3}$ |  | 3 Estimat |  | 95\% CL | 95\% CL |  | 3 | Distrib. ${ }^{3}$ |
| Separable model : F by year |  |  |  |  |  |  |  |  |
| 1 | 2002 | 0.2894 | 9 | 0.2424 | 0.3455 | 0.2644 | 0.3168 | 0.2906 |
| 2 | 2003 | 0.2946 | 8 | 0.2471 | 0.3513 | 0.2693 | 0.3223 | 0.2958 |
| 3 | 2004 | 0.3343 | 9 | 0.2790 | 0.4006 | 0.3048 | 0.3667 | 0.3358 |
| 4 | 2005 | 0.4239 | 9 | 0.3495 | 0.5143 | 0.3842 | 0.4679 | 0.4260 |
| 5 | 2006 | 0.4032 | 11 | 0.3228 | 0.5036 | 0.3599 | 0.4516 | 0.4058 |
| Separable Model: Selection (S) by age |  |  |  |  |  |  |  |  |
| 6 | 0 | 0.1498 | 29 | 0.0841 | 0.2668 | 0.1116 | 0.2011 | 0.1564 |
| 7 | 1 | 0.2262 | 28 | 0.1289 | 0.3970 | 0.1698 | 0.3014 | 0.2357 |
| 8 | 2 | 0.4747 | 8 | 0.3995 | 0.5640 | 0.4347 | 0.5183 | 0.4765 |
| 9 | 3 | 0.6970 | 8 | 0.5882 | 0.8260 | 0.6392 | 0.7601 | 0.6997 |
| 1.0000 Fixed : Reference Age |  |  |  |  |  |  |  |  |
| 10 | 5 | 1.1002 | 9 | 0.9213 | 1.3137 | 1.0050 | 1.2044 | 1.1047 |
| 11 | 6 | 1.1124 | 9 | 0.9179 | 1.3482 | 1.0085 | 1.2271 | 1.1178 |
| 12 | 7 | 1.0846 | 11 | 0.8718 | 1.3493 | 0.9702 | 1.2124 | 1.0913 |
|  | 8 | 1.0000 |  | xed : Las | t true age |  |  |  |
| Separable model: Populations in year 2006 |  |  |  |  |  |  |  |  |
| 13 | 0 | 27777002 | 18 | 19358098 | 39857315 | 23103288 | 33396191 | 28252427 |
| 14 | 1 | 6043461 | 14 | 4588139 | 7960400 | 5250979 | 6955546 | 6103460 |
| 15 | 2 | 2638453 | 10 | 2156853 | 3227589 | 2380630 | 2924199 | 2652439 |
| 16 | 3 | 1349652 | 9 | 1122121 | 1623320 | 1228323 | 1482966 | 1355654 |
| 17 | 4 | 1465554 | 9 | 1227228 | 1750163 | 1338681 | 1604452 | 1471575 |
| 18 | 5 | 2773018 | 9 | 2303524 | 3338203 | 2522614 | 3048279 | 2785465 |
| 19 | 6 | 761340 | 10 | 617967 | 937977 | 684459 | 846857 | 765666 |
| 20 | 7 | 870024 | 12 | 678696 | 1115289 | 766483 | 987552 | 877036 |
| 21 | 8 | 192046 | 14 | 143216 | 257526 | 165348 | 223056 | 194210 |
| Separable model: Populations at age |  |  |  |  |  |  |  |  |
| 22 | 2002 | 137602 | 21 | 90255 | 209787 | 110963 | 170636 | 140824 |
| 23 | 2003 | 136109 | 16 | 97759 | 189503 | 114962 | 161145 | 138063 |
| 24 | 2004 | 279640 | 14 | 209278 | 373660 | 241200 | 324207 | 282715 |
| 25 | 2005 | 181649 | 13 | 138182 | 238789 | 157990 | 208850 | 183426 |
| Recruitment in year 2007 |  |  |  |  |  |  |  |  |
| 26 | 2006 | 11923986 | 26 | 7092988 | 20045354 | 9147941 | 15542453 | 12350180 |
| SSB Index catchabilities |  |  |  |  |  |  |  |  |
| Power model fitted. Slopes (Q) and exponents (K) at age |  |  |  |  |  |  |  |  |
| 27 | 1 Q | 3.150 | 11 | . 649 4 | 4.1012 | . 949 | 3.685 | 3.317 |
| 28 | 1 K | .1162E-04 |  | 1693E-04 | 2620E-04 | 1884E-04 | .2355E-04 | 2227E-04 |

Age-structured index catchabilities
Acoustic survey 1-9+ wr
Linear model fitted. Slopes at age :

| 29 | 1 | $Q$ | 1.181 | 8 | 1.088 | 1.520 | 1.181 | 1.401 | 1.291 |
| :--- | :--- | :--- | :--- | ---: | :--- | :--- | :--- | :--- | :--- |
| 30 | 2 | $Q$ | 1.556 | 6 | 1.465 | 1.873 | 1.556 | 1.763 | 1.660 |
| 31 | 3 | $Q$ | 1.810 | 11 | 1.618 | 2.559 | 1.810 | 2.287 | 2.048 |
| 32 | 4 | $Q$ | 1.840 | 15 | 1.590 | 2.886 | 1.840 | 2.494 | 2.167 |
| 33 | 5 | $Q$ | 1.906 | 16 | 1.633 | 3.066 | 1.906 | 2.628 | 2.267 |
| 34 | 6 | $Q$ | 1.875 | 17 | 1.591 | 3.107 | 1.875 | 2.638 | 2.256 |
| 35 | 7 | $Q$ | 1.738 | 18 | 1.458 | 2.990 | 1.738 | 2.508 | 2.123 |
| 36 | 8 | $Q$ | 1.937 | 18 | 1.623 | 3.338 | 1.937 | 2.798 | 2.367 |
| 37 | 9 | $Q$ | 5.295 | 21 | 4.307 | 10.01 | 5.295 | 8.144 | 6.721 |

Table 2.6.6 (cont) North Sea herring. Model fit parameters, residuals and diagnostics.
IBTS1: 1-5+ wr

| Linear model fitted. Slopes at age : |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 38 | 1 | $Q$ | $.1631 \mathrm{E}-03$ | 6 | $.1537 \mathrm{E}-03$ | $.1961 \mathrm{E}-03$ | $.1631 \mathrm{E}-03$ | $.1847 \mathrm{E}-03$ | $.1739 \mathrm{E}-03$ |
| 39 | 2 | Q | $.1553 \mathrm{E}-03$ | 7 | $.1439 \mathrm{E}-03$ | $.1964 \mathrm{E}-03$ | $.1553 \mathrm{E}-03$ | $.1820 \mathrm{E}-03$ | $.1687 \mathrm{E}-03$ |
| 40 | 3 | $Q$ | $.1093 \mathrm{E}-03$ | 41 | $.7339 \mathrm{E}-04$ | $.3729 \mathrm{E}-03$ | $.1093 \mathrm{E}-03$ | $.2504 \mathrm{E}-03$ | $.1803 \mathrm{E}-03$ |
| 41 | 4 | Q | $.6292 \mathrm{E}-04$ | 41 | $.4226 \mathrm{E}-04$ | $.2147 \mathrm{E}-03$ | $.6292 \mathrm{E}-04$ | $.1442 \mathrm{E}-03$ | $.1038 \mathrm{E}-03$ |
| 42 | 5 | Q | $.3373 \mathrm{E}-04$ | 41 | $.2265 \mathrm{E}-04$ | $.1152 \mathrm{E}-03$ | $.3373 \mathrm{E}-04$ | $.7733 \mathrm{E}-04$ | $.5566 \mathrm{E}-04$ |

MIK 0-wr
Linear model fitted. Slopes at age :
$430 \quad \mathrm{Q} .3562 \mathrm{E}-05 \quad 6.3333 \mathrm{E}-05.4373 \mathrm{E}-05$. 3562E-05 . $4092 \mathrm{E}-05$. $3827 \mathrm{E}-05$

| Parameters of the stock-recruit relationship |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 44 | 1 | a | $.6034 \mathrm{E}+08$ | 21 | $.4900 \mathrm{E}+08$ | $.1147 \mathrm{E}+09$ | $.6034 \mathrm{E}+08$ | $.9312 \mathrm{E}+08$ | $.7675 \mathrm{E}+08$ |
| 45 | 1 | b | $.4209 \mathrm{E}+06$ | 44 | $.2736 \mathrm{E}+06$ | $.1588 \mathrm{E}+07$ | $.4209 \mathrm{E}+06$ | $.1032 \mathrm{E}+07$ | $.7289 \mathrm{E}+06$ |

## RESIDUALS ABOUT THE MODEL FIT

| Separable Model Residuals |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 2002 | 2003 | 2004 | 2005 | 2006 |
| 0 | -0.1458 | -0.3201 | 0.0227 | 0.3945 | -0.1631 |
| 1 | -0.4474 | 0.3089 | -0.3814 | 0.4421 | -0.4124 |
| 2 | 0.0796 | -0.0371 | -0.0782 | 0.0080 | 0.0054 |
| 3 | 0.0686 | 0.0095 | 0.0003 | -0.1432 | 0.0311 |
| 4 | 0.0028 | -0.0615 | -0.0331 | -0.1586 | 0.0007 |
| 5 | -0.2178 | 0.1181 | 0.0414 | 0.0047 | 0.0503 |
| 6 | 0.0243 | -0.0907 | 0.0581 | 0.0401 | -0.0415 |
| 7 | -0.0788 | 0.1050 | 0.2256 | -0.0269 | -0.1751 |
| 8 | 0.0484 | 0.1284 | -0.1489 | 0.3875 | 0.0354 |

SPAWNING BIOMASS INDEX RESIDUALS

| MLAI |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| 1 | -0.2462 | 3275 | 6124 | 6641 | 0.8061 | 0.6382 | 0.7375 | 0.1261 |




Table 2.6.6 (cont) North Sea herring. Model fit parameters, residuals and diagnostics.

| MLAI |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 1 | 0.2369 | 0.1840 | 1708 | 5818 | 0.0850 | 3053 | 0.5061 | 0.6391 |


|  | MLAI |  |
| :---: | :---: | :---: |
|  |  |  |
|  | 2005 | 2006 |
| 1 | 0.2397 | 0.0701 |

AGE-STRUCTURED INDEX RESIDUALS

| Acoustic survey 1-9+ wr |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | * | ******* | ***** | ***** | ******* | ******* | ***** |  |
| 2 | -0.365 | -0.063 | -0.078 | 0.142 | 0.174 | -0.153 | 0.107 | 0.392 |
| 3 | -0.387 | -0.038 | -0.215 | -0.099 | 0.022 | -0.306 | 0.281 | 0.431 |
| 4 | -0.445 | 0.049 | -0.078 | 0.040 | 0.233 | -0.472 | 0.401 | 0.400 |
| 5 | -0.163 | -0.035 | -0.023 | -0.054 | 0.317 | 0.073 | 0.167 | 0.349 |
| 6 | -0.090 | 0.291 | -0.183 | 0.312 | 0.315 | 0.285 | 0.107 | -0.279 |
| 7 | -0.162 | 0.476 | 0.377 | 0.100 | 0.478 | 0.198 | 0.272 | -0.290 |
| 8 | 0.039 | 0.667 | 0.358 | 0.336 | 0.169 | 0.414 | -0.091 | 0.813 |
| 9 | -0.853 | -0.584 | -0.582 | -0.241 | -0.537 | -0.699 | -0.486 | 1.716 |


| Acoustic survey 1-9+ wr |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 1 | -0.193 | -0.330 | -0.248 | 0.436 | -0.297 | 0.075 | 0.303 | 0.194 |
| 2 | 0.519 | -0.175 | -0.142 | -0.312 | 0.179 | -0.188 | 0.325 | -0.298 |
| 3 | 0.422 | 0.108 | 0.072 | -0.102 | 0.109 | 0.038 | -0.378 | -0.114 |
| 4 | 0.390 | 0.383 | -0.177 | 0.231 | -0.027 | -0.261 | -0.255 | -0.325 |
| 5 | 0.309 | 0.637 | -0.324 | 0.229 | 0.086 | -0.248 | -0.610 | -0.336 |
| 6 | 0.567 | 0.727 | 0.139 | 0.095 | -0.049 | 0.067 | -0.281 | -0.899 |
| 7 | -0.596 | 0.925 | 0.318 | 0.574 | -0.492 | -0.154 | -0.033 | -0.171 |
| 8 | -0.064 | -0.213 | 0.170 | 0.599 | -0.078 | -0.575 | -0.377 | -0.221 |
| 9 | 0.835 | 1.259 | 1.241 | 1.161 | 0.304 | -0.037 | 0.115 | -0.969 |


|  | Acoustic survey |  |
| :---: | :---: | :---: |
| Age | 2005 | 2006 |
| 1 | -0.494 | 0.555 |
| 2 | -0.247 | 0.186 |
| 3 | 0.096 | 0.063 |
| 4 | -0.109 | 0.026 |
| 5 | -0.434 | 0.063 |
| 6 | -0.585 | -0.535 |
| 7 | -1.122 | -0.694 |
| 8 | -0.735 | -1.208 |
| 9 | -1.886 | 0.247 |

Table 2.6.6 (cont) North Sea herring. Model fit parameters, residuals and diagnostics.

| IBTS1: 1-5+ wr |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 1 | -0.345 | -0.026 | -0.352 | -0.228 | 0.195 | 0.219 | -0.483 | -0.292 |
| 2 | -1.469 | 0.105 | 0.341 | -0.128 | 1.224 | 0.101 | -0.014 | 0.520 |
| 3 | -0.823 | 0.086 | 0.202 | -0.460 | 0.768 | -0.284 | -0.006 | 0.431 |
| 4 | -0.092 | 0.006 | 0.213 | -0.065 | 0.097 | -0.254 | 0.643 | 0.881 |
| 5 | -0.143 | 0.689 | 0.531 | 0.669 | -0.144 | -1.168 | 0.460 | 0.781 |


| IBTS1: 1-5+ wr |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 1 | -0.173 | 0.186 | -0.011 | -0.031 | 0.204 | 0.589 | 0.483 | -0.625 |
| 2 | -0.070 | 0.734 | 0.730 | 1.012 | -0.689 | 0.108 | -0.093 | 0.003 |
| 3 | 0.446 | 0.476 | 0.741 | 0.643 | -1.509 | 0.080 | -0.616 | 0.374 |
| 4 | 0.089 | 0.258 | 0.719 | 0.456 | -1.115 | -0.350 | -0.830 | 0.559 |
| 5 | 0.362 | 0.396 | 0.444 | -1.051 | -0.380 | 0.166 | -0.167 | 0.354 |


| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.064 | 0.221 | -0.133 | 0.561 | 0.050 | -0.114 | 0.069 | -0.024 |
| 2 | -0.919 | -0.083 | -0.072 | -0.057 | -0.213 | -0.417 | 1.335 | -1.986 |
| 3 | -0.158 | 0.427 | -0.517 | 0.640 | 0.039 | 0.365 | 0.866 | -2.209 |
| 4 | -0.608 | 0.419 | -1.068 | 0.499 | -0.094 | 0.343 | 1.067 | -1.771 |
| 5 | -1.304 | 0.263 | -1.005 | -0.115 | 0.379 | 0.156 | 0.846 | -1.013 |

MIK 0-wr

| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.0643 | 0.2326 | -0.0268 | 0.0133 | -0.3811 | 0.5325 | -0.4719 | 0.1213 |


| MIK 0-wr |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 0 | 0.0943 | 2860 | 0.5000 | 0643 | 3987 | 0.1159 | 0419 | 0.0000 |

Table 2.6.6 (cont) North Sea herring. Model fit parameters, residuals and diagnostics.


## PARAMETERS OF DISTRIBUTIONS OF THE SSB INDICES

DISTRIBUTION STATISTICS FOR MLAI

Power catchability relationship assumed
Last age is a plus-group

| Variance | 0.1012 |
| :--- | ---: |
| Skewness test stat. | 0.3945 |
| Kurtosis test statistic | -0.8273 |
| Partial chi-square | 1.4846 |
| Significance in fit | 0.0000 |
| Number of observations | 34 |
| Degrees of freedom | 32 |
| Weight in the analysis | 0.6000 |

PARAMETERS OF THE DISTRIBUTION OF THE AGE-STRUCTURED INDICES

DISTRIBUTION STATISTICS FOR Acoustic survey 1-9+ wr

Linear catchability relationship assumed


DISTRIBUTION STATISTICS FOR IBTS1: 1-5+ wr

Linear catchability relationship assumed

| Age | 1 | 2 | 3 | 4 | 5 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Variance | 0.0446 | 0.1641 | 0.0055 | 0.0047 | 0.0044 |
| Skewness test stat. | 0.1515 | -1.1821 | -2.7863 | -1.6362 | -1.4818 |
| Kurtosis test statisti | -0.3327 | 0.7375 | 1.7334 | 0.2593 | -0.7119 |
| Partial chi-square | 0.1364 | 0.6158 | 0.0246 | 0.0269 | 0.0283 |
| Significance in fit | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| Number of observations | 24 | 24 | 24 | 24 | 24 |
| Degrees of freedom | 23 | 23 | 23 | 23 | 23 |
| Weight in the analysis | 0.4700 | 0.2800 | 0.0100 | 0.0100 | 0.0100 |

Table 2.6.6 (cont) North Sea herring. Model fit parameters, residuals and diagnostics.

| DISTRIBUTION STATISTICS FOR MIK |  |
| :--- | ---: |
|  |  |
| Linear catchability relationship |  |
|  |  |
| Age | 0 |
| Variance | 0.0523 |
| Skewness test stat. | 0.2141 |
| Kurtosis test statisti | -0.3484 |
| Partial chi-square | 0.1700 |
| Significance in fit | 0.0000 |
| Number of observations | 16 |
| Degrees of freedom | 15 |
| Weight in the analysis | 0.6300 |

ANALYSIS OF VARIANCE

Unweighted Statistics

| Variance |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | SSQ | Data | Parameters | d.f. | Variance |
| Total for model | 109.2433 | 415 | 45 | 370 | 0.2953 |
| Catches at age | 1.5562 | 45 | 25 | 20 | 0.0778 |
| SSB Indices |  |  |  |  |  |
| MLAI | 5.3996 | 34 | 2 | 32 | 0.1687 |
| Aged Indices |  |  |  |  |  |
| Acoustic survey 1-9+ wr | 33.5959 | 154 | 9 | 145 | 0.2317 |
| IBTS1: 1-5+ wr | 49.1683 | 120 | 5 | 115 | 0.4276 |
| MIK 0-wr | 1.2444 | 16 | 1 | 15 | 0.0830 |
| Stock-recruit model | 18.2789 | 46 | 2 | 44 | 0.4154 |


|  | SSQ | Data | Parameters | d.f. | Variance |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Total for model | 5.9548 | 415 | 45 | 370 | 0.0161 |
| Catches at age | 0.7246 | 45 | 25 | 20 | 0.0362 |
| SSB Indices |  |  |  |  |  |
| MLAI | 1.9438 | 34 | 2 | 32 | 0.0607 |
| Aged Indices |  |  |  |  |  |
| Acoustic survey 1-9+ wr | 1.0669 | 154 | 9 | 145 | 0.0074 |
| IBTS1: 1-5+ wr | 1.5428 | 120 | 5 | 115 | 0.0134 |
| MIK 0-wr | 0.4939 | 16 | 1 | 15 | 0.0329 |
| Stock-recruit model | 0.1828 | 46 | 2 | 44 | 0.0042 |

Table 2.7.1. Input to short term prediction.



Table 2.7.2. Management options for North Sea herring.

## Intermediate year (2007) with catch constraint.

| F1 | F2 | F3 | F4 | $\mathbf{F}_{\mathbf{0} \mathbf{1}}$ | $\mathbf{F}_{2-6}$ | $\mathbf{C 1}$ | $\mathbf{C 2}$ | $\mathbf{C 3}$ | C4 | SSB2007 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.328 | 0.042 | 0.012 | 0.007 | 0.064 | 0.336 | 374.9 | 10.5 | 11.6 | 3.4 | 968.7 |

## Prediction year (2008)

| F-values by fleet and total |  |  |  |  | Catches by FLeet |  |  |  |  |  |  |
| :--- | :---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| F1 | F2 | F3 | F4 | $\mathrm{F}_{0-1}$ | $\mathrm{~F}_{2-6}$ | C1 | C2 | C3 | C4 | SSB2008 | SSB2009 |

1. Following management rule, and with catches in IIIa according to a $15 \%$ reduction in WBSS catch

| 0.156 | 0.028 | 0.024 | 0.022 | 0.076 | 0.168 | 171.9 | 8.7 | 14.4 | 6.9 | 1025.3 | 995.5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

2. Following management rule, and with catches in IIIa according to F0.1 for WBSS

| 0.158 | 0.050 | 0.013 | 0.012 | 0.076 | 0.167 | 174.1 | 15.5 | 7.6 | 3.7 | 1026.0 | 999.8 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

3. $15 \%$ reduction in TAC by fleet A , and with catches in IIIa according to a $15 \%$ reduction in WBSS catch

| 0.280 | 0.028 | 0.024 | 0.022 | 0.078 | 0.292 | 289.9 | 8.7 | 14.4 | 6.9 | 944.2 | 821.9 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

4. $15 \%$ reduction in TAC by fleet A, and with catches in IIIa according to F0.1 for WBSS

| 0.280 | 0.050 | 0.013 | 0.012 | 0.077 | 0.289 | 289.9 | 15.4 | 7.6 | 3.7 | 946.3 | 829.0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |  |
| 5. As 1, but with transfer of 3820 tonnes from C-fleet to A-fleet |  |  |  |  |  |  |  |  |  |  |  |
| 0.160 | 0.028 | 0.023 | 0.022 | 0.075 | 0.172 | 175.7 | 8.7 | 13.7 | 6.9 | 1022.9 | 990.4 |
|  |  |  |  |  |  |  |  |  |  |  |  |
| 6. No fishing |  |  |  |  |  |  |  |  |  |  |  |
| 0.0 | 0.0 | 0.00 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 | 0 | 1147.4 | 1309.9 |

## Herring catches 2006, 1st Quarter



Figure 2.1.1: Herring catches (in tonnes)in the North Sea and Division IIIa in 2006 by statistical rectangle. Working group estimates (if available). a.: 1st quarter

Herring catches 2006, 2nd Quarter


Figure 2.1.1: Herring catches (in tonnes) in the North and Division IIIa Sea in 2006 by statistical rectangle. Working group estimates (if available). b.: 2nd quarter

Herring catches 2006, 3rd Quarter


Figure 2.1.1: Herring catches (in tonnes)in the North Sea and Division IIIa in 2006 by statistical rectangle. Working group estimates (if available). c.: 3rd quarter


Figure 2.1.1.: Herring catches (in tonnes)in the North Sea and Division IIIa in 2006 by statistical rectangle. Working group estimates (if available). d.: 4th quarter

Herring catches 2005, All Quarters


Figure 2.1.1: Herring catches (in tonnes) in the North and Division IIIa Sea in 2006 by statistical rectangle. Working group estimates (if available). e: all quarters. Note the wrong heading: figure show catches in 2006 and not 2005.


Figure 2.2.1: Proportions of age groups (numbers) in the total catch of herring in the North Sea (upper, 1960-2006, and middle panel, 1980-2006), and in the total catch of North Sea autumn spawners in 2006 (lower panel).


Figure 2.2.2: Mean vertebrae counts of 2 (upper number), 3 (middle) and $4+$ herring (lower) in the North Sea and Div. IIIa as obtained by Norwegian sampling in the $2^{\text {nd }}$ and $3^{\text {rd }}$ quarter 2006. The transfer area (Western Baltic spring spawners transferred to the assessment of IIIa herring) is indicated.


Figure 2.3.1.1: Survey area coverage in the ICES Coordinated herring acoustic surveys in JuneJuly 2006, by rectangle and nation (WSC = West of Scotland charter vessel; SCO = Scotia; NOR = Johan Hjort; DK = Dana; NL = Tridens; GER = Solea).

D8 D9 E0 E1 E2 E3 E4 E5 E6 E7 E8 E9 F0 F1 F2 F3 F4 F5 F6 F7 F8 F9 G0 G1 G2 G3


Figure 2.3.1.2: Abundance of Autumn spawning herring 1-9+ from combined acoustic survey June-July 2006. Numbers (millions) (upper figure) and biomass (thousands of tonnes) (lower figure).

D8 D9 E0 E1 E2 E3 E4 E5 E6 E7 E8 E9 F0 F1 F2 F3 F4 F5 F6 F7 F8 F9 G0 G1 G2 G3


Figure 2.3.1.3: Numbers (millions) of Autumn spawning herring from combined acoustic survey June-July 2006. 1 ring (upper figure), 2 ring (centre figure), $3+$ (lower figure).

D8 D9 E0 E1 E2 E3 E4 E5 E6 E7 E8 E9 F0 F1 F2 F3 F4 F5 F6 F7 F8 F9 G0 G1 G2 G3


Figure 2.3.1.4: Mean weight and maturity of autumn spawning herring from combined acoustic survey June-July 2006. Four values per ICES rectangle, percentage mature of 2 ring (lower left) and 3 ring fish (lower right), mean weights gram of 1 ring (upper left) and 2 ring fish (upper right), 0 indicates measured percentage mature, + indicates surveyed with zero abundance, blank indicates an unsurveyed rectangle.


Figure 2.3.1.5: Biomass of mature autumn spawning herring from combined acoustic survey in June - July 2006.


Figure 2.3.1.6: Biomass of immature autumn spawning herring from combined acoustic survey in June - July 2006.


Figure 2.3.2.1: North Sea autumn spawners. Orkney/Shetlands 01-15 September 2006. Abundance of larvae $<10 \mathrm{~mm}\left(\mathbf{n} / \mathrm{m}^{2}\right)$


Figure 2.3.2.2: North Sea autumn spawners. Orkney/Shetlands 16-30 September 2006. Abundance of larvae $<10 \mathrm{~mm}$ ( $\mathbf{n} / \mathrm{m}^{2}$ )


Figure 2.3.2.3: North Sea autumn spawners. Buchan 16-30 September 2006. Abundance of larvae $<10 \mathrm{~mm}\left(\mathrm{n} / \mathrm{m}^{2}\right)$


Figure 2.3.2.4: North Sea autumn spawners. Central North Sea 16-30 September 2006. Abundance of larvae $<\mathbf{1 0} \mathbf{~ m m}\left(\mathbf{n} / \mathbf{m}^{2}\right)$


Figure 2.3.2.5: North Sea autumn spawners. Southern North Sea 16-31 December 2006. Abundance of larvae < 11 mm ( $\mathrm{n} / \mathrm{m}^{2}$ )


Figure 2.3.2.6: North Sea autumn spawners. Southern North Sea 1-15 January 2007. Abundance of larvae $<11 \mathrm{~mm}\left(\mathbf{n} / \mathrm{m}^{2}\right)$


Figure 2.3.2.7: North Sea autumn spawners. Southern North Sea 16-31 January 2007. Abundance of larvae $<11 \mathrm{~mm}\left(\mathbf{n} / \mathrm{m}^{2}\right)$


Figure 2.3.2.8: North Sea autumn spawners. Larval Abundance Index time-series for a collection of areas and sampling periods (Orkney/Shetlands 2nd half of September top left panel, Buchan 2nd half of September top right, central North Sea lower left, southern North Sea lower right. Due to historic reasons the abundance in the CNS is given as the mean of three surveys and in the SNS as the sum of three).


Figure 2.3.2.9: North Sea autumn spawners. Comparison of spawning stock size estimates from the Herring Assessment Working Group (ICES, 2006; bold line) and the year effects fitted to the larval abundances in the multiplicative model (symbols with error bars). The MLAI estimates have been rescaled to the mean of the WG estimates. Error bars indicate +/- one standard error of larval survey abundance estimates. Note the $\log y$ axis.

 Longitude

1-ringers Yearclass 2004

$-3-2-1061243456789101112$ Longitude

1-ringers Yearclass 2005

$\begin{array}{lllllllllllll}-3 & -2 & -1 & 0 & 1 & 2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 \\ 1011\end{array}$ Longitude

Figure 2.3.3.1. North Sea herring. Distribution of 1-ringer herring, year classes 2003-2005. Abundance estimates of 1-ringers within each statistical rectangle are based on

## Mean length 1-ringers from IBTS 2007



Figure 2.3.3.2. North Sea herring. Mean length (mm) of 1-ringer herring caught during IBTS ${ }^{\text {st }}$ Quarter


Figure 2.3.3.3. North Sea herring. Distribution of 0-ringer herring, year classes 2004-2006. Abundance estimates of 0-ringers within each statistical rectangle are based on MIK catches during IBTS in February 2005-2007. Areas of filled circles illustrate densities in no $\mathbf{m}^{-2}$, the area of a circle extending to the border of a rectangle represents $\mathbf{1} \mathbf{m}^{-2}$


Figure 2.3.3.4 North Sea herring. Absolute (no ${ }^{*} \mathbf{1 0}^{\mathbf{9}}$ ) and relative abundance of 0 -ringers in the area west of $2^{\circ} \mathrm{E}$ in the North Sea. Abundances are based on MIK sampling during IBTS, the relative abundance in the western part is estimated as the number of 0 -ringers west of $2^{\circ} \mathrm{E}$ relative to total number of 0 -ringers.


Figure 2.5.1 North Sea herring. Relationship between indices of 0 -ringers and $\mathbf{1}$-ringers for year classes 1977 to 2005. The 2005 relation is shown as a filled square, the present 0 -ringer index for year class 2006 is indicated by an arrow.

Time series of recruitment indices


Figure 2.5.2 North Sea herring. Time series of 0-ringer and 1-ringer indices. Year classes 1976 to 2006 for 0-ringers, year classes 1977-2005 for 1-ringers.


Figure 2.5.3. North Sea herring. Trend in recruitment of 1-ringers from year class 1958 to 2005. Data from the 2007 ICA assessment of the North Sea autumn spawned herring.


Figure 2.6.1. North Sea herring. Comparison of mean reference relationship between mean F and SSB for:

- Assessment point estimate using catch and all indices using the benchmarked procedure from last year
- Variance - Covariance (V-CV) uncertainty of F and SSB using bootstrap estimates
- Flat selection of $\mathbf{F}$ at age $7,8,9+$
- Each individual fleet as the only tuning indices (Acoustic 1-9+wr, IBTS 1-5+wr, MIK 0wr and MLAI SSB index)


Figure 2.6.2 North Sea herring. Comparison of weighted residuals for assessment 2006 (left panels) and 2007 (right panels) using same procedure. Dark bubbles represent residual values greater than 0 , white bubbles less then 0 , left hand dark bubble represents a scaling value of $\mathbf{1 . 0}$. Plot of (a) catch residuals at age for the separable period; (b) MIK 0 group index; (c) MLAI SSB index; (d) Acoustic survey with age 1 only from 1997 onwards; (e) IBTS survey.


Figure 2.6.3. North Sea herring. Retrospective ICA plots for SSB, mean F on ages 2-6, and recruitment.


Figure 2.6.4. North Sea herring. Retrospective ICA plots for SSB, mean $F$ on ages 2-6, and recruitment based on an assessment using only MIK and Acoustic survey data.


Figure 2.6.5. North Sea herring. Retrospective ICA plots for SSB, mean F on ages 2-6, and recruitment based on an assessment using only MIK and IBTS survey data.


Figure 2.6.6. North Sea herring. Retrospective ICA plots for SSB, mean $F$ on ages 2-6, and recruitment based on an assessment using only MIK and MLAI survey data.


Figure 2.6.7. North Sea herring. Stock summary according to the final ICA assessment: SSB, mean F on ages 2-6 and ages 0-1, and recruitment. The reference line for SSB corresponds to 800000 tonnes (Blim), while the reference lines for mean $F$ correspond to 0.25 (solid line) and 0.12 (dashed line).


Figure 2.6.8. North Sea herring. Diagnostics of selection pattern from the final ICA assessment. Top left: bubbles plot of log catch residuals by age (weighting applied) and year ( 5 yr separable period). Top right: estimated selection parameters (relative to $4 \mathbf{w r}$ ) with $95 \%$ confidence intervals. Middle left: marginal totals of log residuals by year. Middle right: marginal totals of log residuals by age (wr).


Figure 2.6.9. North Sea herring. Diagnostics of MLAI survey catchability from the final ICA assessment. Top left: VPA estimates of SSB (line) and SSB predictions made from index observations with 95\% confidence intervals. Top right: scatterplot of index observations versus VPA estimates of SSB with the best-fit catchability model (power function). Middle left: log residuals of catchability model by VPA estimate of SSB. Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of $\log$ residuals.


Figure 2.6.10. North Sea herring. Diagnostics of Acoustic survey catchability at $1 \mathbf{w r}$ from the final ICA assessment. Top left: VPA estimates of numbers at 1 wr (line) and numbers predicted from index abundance at 1 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 1 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 1 wr. Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of $\log$ residuals. Bottom right: autocorrelogram of $\log$ residuals.


Figure 2.6.11. North Sea herring. Diagnostics of Acoustic survey catchability at 2 wr from the final ICA assessment. Top left: VPA estimates of numbers at 2 wr (line) and numbers predicted from index abundance at 2 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 2 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 2 wr. Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of $\log$ residuals.


Figure 2.6.12. North Sea herring. Diagnostics of Acoustic survey catchability at $3 \mathbf{w r}$ from the final ICA assessment. Top left: VPA estimates of numbers at 3 wr (line) and numbers predicted from index abundance at 3 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 3 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 3 wr . Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.13. North Sea herring. Diagnostics of Acoustic survey catchability at 4 wr from the final ICA assessment. Top left: VPA estimates of numbers at 4 wr (line) and numbers predicted from index abundance at 4 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 4 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 4 wr . Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.14. North Sea herring. Diagnostics of Acoustic survey catchability at 5 wr from the final ICA assessment. Top left: VPA estimates of numbers at 5 wr (line) and numbers predicted from index abundance at 5 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 5 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 5 wr . Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.15. North Sea herring. Diagnostics of Acoustic survey catchability at $6 \mathbf{w r}$ from the final ICA assessment. Top left: VPA estimates of numbers at 6 wr (line) and numbers predicted from index abundance at 6 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 6 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at $6 \mathbf{w r}$. Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of $\log$ residuals.


Figure 2.6.16. North Sea herring. Diagnostics of Acoustic survey catchability at 7 wr from the final ICA assessment. Top left: VPA estimates of numbers at 7 wr (line) and numbers predicted from index abundance at 7 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 7 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 7 wr . Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of $\log$ residuals.


Figure 2.6.17. North Sea herring. Diagnostics of Acoustic survey catchability at 8 wr from the final ICA assessment. Top left: VPA estimates of numbers at 8 wr (line) and numbers predicted from index abundance at 8 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 8 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 8 wr . Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.18. North Sea herring. Diagnostics of Acoustic survey catchability at $9+\mathbf{w r}$ from the final ICA assessment. Top left: VPA estimates of numbers at 9+ wr (line) and numbers predicted from index abundance at $9+$ wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at $9+\mathbf{w r}$ with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at $9+\mathbf{w r}$. Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.19. North Sea herring. Diagnostics of IBTS survey catchability at $1 \mathbf{w r}$ from the final ICA assessment. Top left: VPA estimates of numbers at $1 \mathbf{w r}$ (line) and numbers predicted from index abundance at 1 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at $1 \mathbf{w r}$ with the best-fit catchability model (linear function). Middle left: $\log$ residuals of catchability model by VPA estimate of numbers at $1 \mathbf{w r}$. Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.20. North Sea herring. Diagnostics of IBTS survey catchability at 2 wr from the final ICA assessment. Top left: VPA estimates of numbers at 2 wr (line) and numbers predicted from index abundance at 2 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 2 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 2 wr . Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.21. North Sea herring. Diagnostics of IBTS survey catchability at 3 wr from the final ICA assessment. Top left: VPA estimates of numbers at 3 wr (line) and numbers predicted from index abundance at 3 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 3 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 3 wr . Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.22. North Sea herring. Diagnostics of IBTS survey catchability at 4 wr from the final ICA assessment. Top left: VPA estimates of numbers at 4 wr (line) and numbers predicted from index abundance at 4 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 4 wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at 4 wr . Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.23. North Sea herring. Diagnostics of IBTS survey catchability at $5+\mathbf{w r}$ from the final ICA assessment. Top left: VPA estimates of numbers at $5+\mathrm{wr}$ (line) and numbers predicted from index abundance at $5^{+}$wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at $5+$ wr with the best-fit catchability model (linear function). Middle left: log residuals of catchability model by VPA estimate of numbers at $5+$ wr. Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.


Figure 2.6.24. North Sea herring. Diagnostics of MIK survey catchability at 0 wr from the final ICA assessment. Top left: VPA estimates of numbers at 0 wr (line) and numbers predicted from index abundance at 0 wr with $95 \%$ confidence intervals. Top right: scatterplot of index observations versus VPA estimates of numbers at 0 wr with the best-fit catchability model (linear function). Middle left: $\log$ residuals of catchability model by VPA estimate of numbers at $\mathbf{0} \mathbf{w r}$. Middle right: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals. Bottom right: autocorrelogram of log residuals.






Figure 2.8.1

Results of medium term predictions for North Sea herring
Scenario 1: Harvest rule with not constraint on year-to-year change in TAC.
Upper panels: Percentiles for SSB, and risk to Blim
Middle panels: Percentiles for catch
Lower panels: Percentiles for fishing mortality
Left: Fleet A
Right: Fleets B-D


Figure 2.8.2
Results of medium term predictions for North Sea herring
Scenario 2: Harvest rule with $15 \%$ constraint on year-to-year change in TAC.
Upper panels: Percentiles for SSB, and risk to Blim
Middle panels: Percentiles for catch
Lower panels: Percentiles for fishing mortality
Left: Fleet A
Right: Fleets B-D



Figure 2.8.3
Results of medium term predictions for North Sea herring
Scenario 3: Fixed intended F0-1 = 0.12 and F2-6 = 0.25
Upper panels: Percentiles for SSB, and risk to Blim
Middle panels: Percentiles for catch
Lower panels: Percentiles for fishing mortality
Left: Fleet A
Right: Fleets B-D


Figure 2.8.4

Results of medium term predictions for North Sea herring
Scenario 4: As scenario 1 (harvest rule with $15 \%$ constraint on year-to-year change in TAC), but with higher assumed CV on recruitment of $\mathbf{0 . 5 8}$.
Upper panels: Percentiles for SSB, and risk to Blim
Middle panels: Percentiles for catch
Lower panels: Percentiles for fishing mortality
Left: Fleet A
Right: Fleets B-D





Figure 2.8.5

Results of medium term predictions for North Sea herring
Scenario 5: As scenario 1 (harvest rule with $15 \%$ constraint on year-to-year change in TAC), but with a lower breakpoint in the stock-recruit function at $\mathbf{5 0 0} \mathbf{0 0 0}$ tonnes
Upper panels: Percentiles for SSB, and risk to Blim
Middle panels: Percentiles for catch
Lower panels: Percentiles for fishing mortality

## Left: Fleet A

Right: Fleets B-D


Figure 2.8.6 Results of medium term predictions for North Sea herring. Effect of implementation error. Risk to Blim in 2017 with and without $15 \%$ constraint on catch variation.




Figure 2.10.1 North Sea herring: historic uncertainty in recruitment, SSB and mean Fages 2-6 wr from 1960-2006 at 5, 25, 50, 75, 95 percentiles estimated by bootstrap of parameter residuals based on variance covariance method in ICA.


Figure 2.10.2 North Sea herring: cohort historic retrospectives for yearclasses currently in terminal year of the assessment, 1997 to 2005.


Figure 2.11.1. North Sea herring. Comparison of TACs for total North Sea and IVc and VIId


Figure 2.11.2. Herring in IVc and VIId. Comparison of historical catches and TACs


Figure 2.11.3 Downs herring. Index (nos per hr) of small ( $<13 \mathrm{~cm}$ ) 1-ringers in the North Sea and proportion of small 1-ringers versus all sizes in the North Sea (from table 2.3.3.3)


Figure 2.11.4. Downs herring. Larval Abundance Index (LAI) in the Channel area (line), calculated as the sum of surveys per year class 1975-2006, and preliminary MIK survey results in the Channel area (early spring 1995-2007, no data available in 1996 and 2001).


Figure 2.11.5 : Acoustic transects in the Eastern Channel. Herring detections are mainly concentrated in front of Boulogne and in less proportions (?) in the southern part of the area.


Figure 2.11.6 Catch composition by age from the pelagics hauls. Age groups 5, 6 and 7 represent respectively $11 \%, 46 \%$ and $18 \%$ of the total.

## 3 Herring in Division IIIa and Subdivisions 22-24 [update assessment]

### 3.1 The Fishery

### 3.1.1 ACFM advice and management applicable to 2006 and 2007

At the ACFM (May) meeting in 2006, it was stated that the status of the stock is unknown relative to safe biological limits, because reference points have not been determined. SSB has been stable or has slightly increased over a number of years. Fishing mortality estimates for 2006 are 0.52 for adults and 0.18 for the juveniles (1-ringers).

ACFM recommended in 2006 that, since the current fishing mortality has lead to a stable or increased SSB, the fishing mortality should not be allowed to increase. This would correspond to catches in 2006 less than 95000 t and less than 99000 t in 2007. According to the recent geographic distribution of catches, approximately half of the total catches should be taken from Subdivisions 22-24.

The EU and Norway agreement on a herring TACs set for 2006 was 81600 t in Division IIIa for the human consumption fleet and a by-catch ceiling of 20528 t to be taken in the small mesh fishery. In 2006 the EU and Norway agreement on herring TACs for 2007 in Division IIIa was 118860 t .

In previous years the International Baltic Sea Fishery Commission (IBSFC) set no special TAC for Subdivisions 22-24. In 2006, a TAC (47500 t) was set for the first time on the Western Baltic stock component. The TAC for 2007 was set at 49500 t .

### 3.1.2 Catches in 2006

Herring caught in Division IIIa are a mixture of North Sea Autumn Spawners (NSAS) and Western Baltic Spring Spawners (WBSS). This Section gives the landings of both NSAS and WBSS, but the stock assessment applies only to the spring spawners.

It is important to note that 2000 tonnes of Danish landings were, on a very early stage lost from the data due to a programming error. This error was not discovered in time and the catch in numbers in the present section therefore does not account for this. However, the error does not influence biological samples of size, age and proportions of spawning type. All relevant tables will be updated with the correct information in next year's report. The eventual effects of the missing catches on the assessment and short term projections for WBSS herring were found to be insignificant for estimates of SSB, F and recruitment.

Landings from 1985 to 2006 are given in Table 3.1.1. In 2006 the total landings in Division IIIa and Subdivisions 22-24 has decreased to 93000 t , which is the second lowest value of the time series (1986-2006), only 2004 was slightly lower. The decrease in landings and their resemblance with 2004 is evident in the catches from the Kattegat and the Skagerrak. The German landings have increased slightly for the last three years in Subdivision 22-24, but are still diminutive in Division IIIa. The overall fishing pattern has changed in the last few years. As in previous years the 2006 landing data are calculated by fleet according to the fleet definitions used when setting TACs.

The fleet definitions used since 1998 are:

- Fleet C: directed fishery for herring in which trawlers (with 32 mm minimum mesh size) and purse seiners participate.
- Fleet D: All fisheries in which trawlers (with mesh sizes less than 32 mm ) and small purse seiners, fishing for sprat along the Swedish coast and in the Swedish fjords, participate. For most of the landings taken by this fleet, herring is landed as by-catch. Danish and Swedish by-catches of herring from the sprat fishery and the Norway pout and blue-whiting fisheries are listed under fleet D.
- Fleet F: Landings from Subdivisions 22-24. Most of the catches are taken in a directed fishery for herring and some as by-catch in a directed sprat fishery.

In Table 3.1.2 the landings are given for 2001 to 2006 in thousands of tonnes by fleet (as defined by HAWG) and quarter.

### 3.2 Biological composition of the catch

Table 3.2.1 and Table 3.2.2 show the total catch (autumn- and spring-spawners combined) in numbers and mean weight-at-age in the catch for herring by quarter and fleet landed from Skagerrak and Kattegat, respectively. The total numbers and mean weights-at-age for herring landed from Subdivisions 22-24 are shown in Table 3.2.3.

The level of sampling of the commercial landings was generally acceptable (Table 3.2.4). In the cases of missing samples the corresponding landings were minor. Where sampling was missing in areas and quarters on national landings, sampling from either other nations or adjacent areas and quarters were used to estimate catch in numbers and mean weight-at-age (Table 3.2.5).

Based on the proportions of spring- and autumn-spawners in the landings (Table 3.2.6 and see Section 3.2.2 for more details) catches were split between NSAS and WBSS.

The total numbers and mean weight-at-age of the WBSS and NSAS landed from Kattegat, Skagerrak, and Div. IIIa respectively was then estimated by quarter and fleet (Table 3.2.73.2.12).

The total catch (SOP) of the WBSS taken in the North Sea + Div. IIIa in 2006 were estimated to be 47070 t , and has thereby increased in the last two years from the very low levels observed in 2004 and 2003 of 35000 and 38000 respectively (Table 3.2.13). However, they are still far below the values observed up to the mid nineties.

Total catches (SOP) of WBSS from the North Sea, Div. IIIa, and Subdivisions 22-24 respectively, by quarter, was estimated for 2006 (Table 3.2.14). Additionally, the total catches of WBSS in numbers and tonnes (SOP), divided between the North Sea + Div. IIIa and Subdivisions 22-24 respectively for 1991-2006, are presented in Tables 3.2.15 and 3.2.16.

Catches (SOP) of WBSS from Subdivisions 22-24 have remained rather stable for the last four years at levels just above 40000 t , which also is the lowest level found in the time series (1991-2006) (Table 3.2.16).

The total catch (SOP) of NSAS in Div. IIIa amounted to 15015 t in 2006, which is only $50 \%$ of the 2005 value, and the lowest ever in the time series (1991-2006). The decrease relative to 2005 was mainly due to a proportionally large drop in representation of the 2004 year class in the catches (Table 3.2.17).

### 3.2.1 Quality of Catch Data and Biological Sampling Data

The amount of discards for 2006 is assumed to be insignificant as in previous years. However, no quantitative estimates of discards were available to the Working Group.
Table 3.2 .4 shows the number of fish aged by country, area, fishery and quarter. The overall sampling in 2006 more than meets the recommended level of one sample per 1000 t landed
per quarter. Coverage of areas, times of the year and gear (mesh size) was acceptable. One exception is a complete lack of samples to cover catches from Subdivision 23 comprising 2 477 t .

### 3.2.2 Stock composition in the catch

Catches of herring in the Kattegat, the Skagerrak and the Eastern part of the North Sea are taken from a mixture of two main spawning stocks. These are 1+ ringers of the Western Baltic Spring Spawners (WBSS) and 0-2-ringers from the North Sea Autumn Spawners (NSAS). The winter spawning Downs herring are included under NSAS (see stock annex 2). An uncertain amount of spring spawners belonging to local spawning populations in the Skagerrak/Kattegat area are likely to contribute to the catches. However due to lack of knowledge concerning these, they are included under WBSS (see also stock annex 2). As in recent years the WG uses the analysis of individual otolith microstructure for determination of spawning type in age-class stratified random sub-samples of herring in Division IIIa (see stock annex 2). The split between WBSS and NSAS in the eastern North Sea is limited to an area also referred to as the transfer area (ICES rectangles: 43F3 to 43F7, 44F3 to 44F6, 45 F 3 to 45F6, 46F3 to 46F6, and 47F3 to 47F6 (see also Figure 2.2.2)), under the assumption that the geographical distribution of WBSS into the North Sea is within the borders of the transfer area.

For the present year the otolith-based method has been exclusively applied for the Division IIIa split. For Subdivisions 22, 23 and 24 it was assumed that all individuals belong to the WBSS stock, even when otolith microstructure indicate occurrence of autumn spawners in the surveys or in samples of commercial catches (see stock annex 2).

Different area based TACs and by-catch ceilings are set for herring in Divisions IIIa and IV. However during summer feeding migrations components of WBSS and NSAS mix in both areas Divisions IIIa and IV East. A recently finalised research project has explored ways to regulate the fishing mortality of NSAS and WBSS individually within Divisions IV and IIIa (IMHERSKA). Results indicate that a set of proposed métiers for the Danish herring fisheries, to some degree, fished selectively with respect to stock (WBSS and NSAS) and fish size, in specific areas and quarters (IMHERSKA final report 2007 in prep.). It is also of note that the results agree with the existing knowledge on migration behaviors of the respective stocks.

### 3.2.2.1 Spring-spawning herring in the North Sea

Catches from the transfer area in the eastern North Sea in 2006 were split by analysis of Norwegian and Danish samples from landings (see Figure 2.2.2 for details about the transfer area). Mean vertebral counts from the Norwegian samples and otolith microstructure readings from the Danish samples were used to estimate the proportion of WBSS. Samples were missing in the $4^{\text {th }}$ quarter for 1 to 3 -ringers and were inferred from neighbouring quarters. The sources of data for splitting between NSAS and WBSS in the transfer area are:

|  | 1-RINGERS | 2-RINGERS | 3-RINGERS | 4+-RINGERS |
| :--- | :--- | :--- | :--- | :--- |
| $\mathbf{1}^{\text {st }}$ <br> quarter | DK samples (landings) | DK \& NOR samples <br> (landings) | DK samples (landings) | DK \& NOR samples <br> (landings) |
| $\mathbf{2}^{\text {nd }}$ <br> quarter | DK samples (acoustic <br> + landings) | DK \& NOR samples <br> (acoustic + landings) | DK \& NOR samples <br> (acoustic + landings) | DK \& NOR samples <br> (acoustic + landings) |
| $\mathbf{3}^{\text {rd }}$ <br> quarter | DK \& NOR samples <br> (acoustic+ landings) | NOR samples <br> (landings) | NOR samples <br> (landings) | NOR samples <br> (landings) |
| 4 <br> quarter | inferred from <br> neighbouring quarters | inferred from <br> neighbouring quarters | inferred from <br> neighbouring quarters | DK samples (landings) |

Resulting proportions of WBSS can be found in Section 2.2.2.

### 3.2.2.2 Autumn spawners in Division Illa

The proportions and the analysed numbers are presented in Table 3.2.6.
For commercial landings in 2006 the split of the Swedish and Danish landings was conducted using the proportion by age in the combined samples of Swedish and Danish microstructure analyses. The estimation of the proportion of spring- and autumn-spawners in the landings from Division IIIa was performed on the basis of 4449 (2903 Danish and 1546 Swedish) otolith microstructure analyses in 2006. Data were disaggregated by area (Kattegat and Skagerrak), quarter ( $1-4$ ) and age group ( $1-8+\mathrm{wr}$ in $1^{\text {st }}$ quarter and $0-8+\mathrm{wr}$ in $2^{\text {nd }}, 3^{\text {rd }}$ and $4^{\text {th }}$ quarter).

Generally, sampling for split in 2006 covered younger age classes (0-2-ringers). In cases where sampling of older age-classes had fewer than 12 individuals per cell (area, quarter, and wr) samples were supplemented with survey samples and/or the cells were pooled to combine age groups (for details see table 3.2.6).

All herring found in subdivisions 22-24 are treated as Western Baltic spring spawners (see stock annex 2).

### 3.2.2.3 Accuracy and precision in stock identification

The stock classification using visual inspection of otolith microstructure has been validated objective criteria as described in a recent publication (Clausen et al. 2007). The correspondence between results from visual inspection by experienced readers and back calculated hatch date from counted microstructures was high, with misclassification levels of $5 \%$ and $3 \%$ for autumn/winter and spring spawners respectively. All of the Danish routine samples for the stock identification are interpreted by experienced readers. However, in the case of spawning type infidelity this validation method would show false misclassification. Therefore, an objective method of hatch time estimation was also employed, counting daily increments in 0 -group herring hatched during different seasons. Visual inspection and objective estimation agreed to $89 \%$, and confusion between autumn and winter spawners was explained by overlapping hatch periods. Older herring have been classified using multiple linear regression of hatch time versus median increment width.

Issues of precision and further development of methods are dealt with in the stock annex.

### 3.3 Fishery Independent Information

### 3.3.1 International Bottom Trawl Survey in Division IIIa

The survey indices were split into spring and autumn spawning components by microstructure analysis of otoliths (section 3.2.2) except for $20013^{\text {rd }}$ quarter and $20021^{\text {st }}$ quarter when vertebrae counting methods were used. The estimates of the abundance by age of the spring spawning component in the Kattegat (SD21) are presented in Table 3.3.1 and Table 3.3.2. The estimated mean value for 1-ringers in $20071^{\text {st }}$ quarter is lower than the average and similar to values observed in 2005 and 2006. The older age classes show a clear decrease with the lowest observed value for age 3 and age 5 and the second lowest for age 4. For $3^{\text {rd }}$ quarter survey indices, the value for 1 -ringers in 2006 is around the average of the time-series while the abundance of 3-ringers is the lowest on record.

### 3.3.2 Summer Acoustic Survey in Division IIIa

The acoustic survey from 23 June to 6 July 2006 covered the area in the Skagerrak and the Kattegat. Details of the survey are given in the 'Report of the Planning Group for Herring Surveys' (ICES 2007/LRC:01). The estimated spawning biomass (3+) of Western Baltic

Spring Spawning herring (WBSS) in 2006 was about 244000 tonnes, showing an increase compared to the previous year of about $105 \%$. The results from this survey are summarised in Table 3.3.3.

### 3.3.3 Autumn Acoustic Survey in Subdivisions 22-24

A joint German-Danish acoustic survey was carried out with R/V "SOLEA" between 5 and 24 October 2006 in the Western Baltic covering Subdivisions 21, 22, 23 and 24. A full survey report is given in the Report of the Planning Group for Herring Surveys (ICES 2007/LRC:01). The results for 2006 are presented in Table 3.3.4. The herring stock was estimated to be about 211000 tonnes in Subdivisions 22-24 (Table 3.3.4). This is an increase of $11 \%$ compared to the last year estimate.

### 3.3.4 Larvae Surveys

Herring larvae surveys in the western Baltic were conducted in weekly intervals during the 2006 spawning season. During the last decade, the Rügen herring larvae surveys in the Greifswalder Bodden aimed at delivering a fishery independent recruitment estimate for the WBSS assessment. The resulting N30 index (extrapolated abundance of larvae at 30 mm length) has shown to reliably predict very strong year classes, however it failed to predict year classes of intermediate strength.

The results for 2006 were not available at the meeting (Table 3.3.5).

### 3.4 Mean weights-at-age and maturity-at-age

Mean weights at age in the catch in the $1^{\text {st }}$ quarter were used as stock weights (Table 3.2.14). The maturity ogive was assumed constant between years. The same maturity ogive was used as in the HAWG 2006:

| W-RINGS | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8 +}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Maturity | 0.00 | 0.00 | 0.20 | 0.75 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 |

### 3.5 Recruitment

Indices of 0-ringer abundance of the Western Baltic Spring Spawning herring (WBSS) in Subdivisions 22-24 for 2006 were available from the autumn acoustic survey in Subdivisions 22-24 (see also Table 3.3.5) while results for 2006 larval survey during spawning time were not available at the meeting. The index of the 0 -ringer in 2006 from the autumn acoustic survey was similar compared to the latest years with a slight increase compared to 2005. The acoustic recruitment indices of the 0 -ringer and 1 -ringer were similar to previous years (Figure 3.5.1). The total number of individuals in the stock from the autumn acoustic survey was also similar to the last year estimates as well as the values for the older age classes.

### 3.6 Assessment of western Baltic spring spawners in Division IIIa and Subdivisions 22-24

### 3.6.1 Input data

Catch in numbers at age from 1991 to 2006 were available for Subdivision IVa (East), Division IIIa and Subdivisions 22-24 (Table 3.6.1) and as proportion at age (Figure 3.6.1). Years before 1991 have been excluded due to lack of reliable data for splitting spawning type and also due to a large change in fishing pattern caused by changes in the German fishing fleets.

Mean weights at age in the landings are found in Table 3.6.2 and in Figure 3.6.2. The proportions of F and M before spawning were assumed constant between years. F-prop was set to be 0.1 and M-prop 0.25 for all age groups. Natural mortality was assumed constant at age and equal to $0.3,0.5$, and 0.2 for 0 - ringers, 1 - ringers, and $2+$ ringers respectively (Table 3.6.4). The estimates of natural mortality were derived as a mean for the years 1977-1995 from the Baltic MSVPA (ICES 1997/J:2).

Available survey indices (Tables 3.3.1-3.3.4) were:
FLT1: Hydroacoustic survey in Division IIIa \& Sub-division IVa East, July 19912006, 0-8+ ringers

FLT2: Hydroacoustic survey in Subdivisions 22, 23 and 24, Oct. 1991-2006, 0-8+ ringers

FLT3: IBTS in Division IIIa, Quarter 1, 1991-2007, 1-5 ringers
FLT4: IBTS in Division IIIa, Quarter 3, 1991-2006, 1-5 ringers
All are age-structured indices. None of the indices covered the total spatial distribution of the WBSS stock and the indices covered the following quarters and areas:

| Survey area | Quarter 1 | QUarter 2 | Quarter 3 | Quarter 4 |
| :---: | :---: | :---: | :---: | :---: |
| Division IIIa | FLT3 | - | FLT1 and FLT4 | - |
| Subdivisions 22-24 | - | - | - | FLT2 |

Subsets of these data series representing selected age groups were constructed to give a better representation of the stock (see section 3.6.3).

### 3.6.2 ICA settings

The following settings (Table 3.6.6) were used in 2007, similar to 2006:

- The period for the separable constraint: 5 years (2002-2006).
- $\quad$ The weighing factor to all indices (lambda $=1$ ).
- A linear catchability model for all indices
- The reference F set at age 4 and the selection $=1$ for the oldest age.
- The catch data were down-weighted to 0.1 for 0 -ringer herring.


### 3.6.3 Exploration by individual survey indices

Given that this is an update assessment only a limited exploration was carried out similar to last year. Exploratory runs of catch data with single indices were performed using the general ICA-setting mentioned above (Section 3.6.2). A summary of the results from these runs is presented in Figures 3.6.3 and 3.6.4.

No larval survey data was available for 2006. The IBTS in Kattegat Q1 (FLT3) indicate a high F of 2.4, somewhat higher than the hydro-acoustic survey indices in Division IIIa (FLT1a and FLT1b) being 1 and 1.5 respectively, whereas the Acoustic survey indices in Subdivisions 2224 (FLT2a and FLT2b) and the IBTS index in Kattegat Q3 (FLT4) suggest low F of 0.3, 0.3 and 0.2 respectively.

With no larval index for 2006 only the only recruitment indices available were 0 -ringer Acoustic in SD 22-24 and 1-ringer Acoustic in SD 22-24. Recent trends in log transformed values of the time series from 1991 show no exceptional development (Figure 3.5.1). The tuning fleet choice and the settings for the final ICA run for the 2006 assessment were therefore the same as in the last two years' assessments with fleets FLT1b, FLT2b, and FLT4. The biological reasoning behind the choice of indices with restricted numbers of age classes is
that there is only a partial migration of age 0-1 ringers to the Division IIIa in the summer and that ages older than 5-ringers are poorly represented in the Subdivision 22-24 acoustic surveys and in the IBTS.

### 3.6.4 Final Assessment

This assessment conforms to an update assessment of WBSS herring, input data (years 19912006, Ages $0-8+$ ringers) are given in the following tables:

- $\quad$ Catch in number (Table 3.6.1)
- Weight in catch (Table 3.6.2)
- Weight in stock (Table 3.6.3)
- Natural mortality (Table 3.6.4)
- Maturity (see text table in section 3.4)

The following surveys were included (Tables 3.6.5a-c):

- FLT 1b: DK Hydroacoustic survey in Division IIIa+ SD IVaE, July 1991-2006, excl. 1999, 2-8+ ringers
- FLT 2b: GER Hydroacoustic survey in Subdivisions 22, 23 and 24, Oct 19912006, 0-5 ringers
- FLT 4: IBTS in Kattegat, Quarter 3, 1991-2006, 1-5 ringers

The final model settings are shown in Table 3.6.6. The output data are given in Tables 3.6.73.6.16. The estimated SSB for 2006 is about 184500 tonnes with a mean fishing mortality (ages 3-6) of 0.52 (Table 3.6.9, Figure 3.6.6). As the previous year, the model diagnostics show a rather well defined minimum SSQ response-curve for all age-indices except age-index 1 (Acoustic Survey in Division IIIa+IVaE) that is somewhat flat (Figure 3.6.5). The minimum SSQ for the Acoustic Survey in Subdivisions 22-24 (age-index 2) finds an intermediate compromise between the high F of age-index 1 (Acoustic Survey in Division IIIa+IVaE) and the low F of age-index 3 (IBTS Kattegat Q3).

The marginal totals of residuals between the catch and the separable model (scrutinised on screen in ICA-view) are overall small, as well as reasonably trend-free in the separable period (2002-2006) (see Figure 3.6.7). However, as already noted in last years assessment the largest residuals and most of the year effects are again caused by 0 -ringers that are down-weighted in the analysis but still appears with full weight in the residual plot of the ICA diagnostics. For values see Table 3.6.12.

The diagnostics for the three surveys does not repeat the trend of low acoustic and high IBTS residuals seen in last year assessment, with values for IBTS in the last two years are now in the same order of magnitude when compared to Acoustic surveys. The Acoustic Survey in Division IIIa+IVaE and the Acoustic Survey in Subdivisions 22-24 showed a mix of negative and positive residuals for 2006 (Figure 3.6.8), with the Acoustic Survey in Subdivisions 22-24 resembling the pattern of IBTS Kattegat Q3 survey. All surveys had noisy fits to population estimates for the younger and older age-classes, and somewhat better for the intermediate ones.

The catch-at-age unweighted variance component is of the same magnitude as the individual acoustic survey variance components (Table 3.6.16), however in the unweighted statistics down-weighting of the 0 -ringers is not accounted for, and this age contribute quite some variation with a C.V. of $57 \%$ compared to about $14-19 \%$ for the $2+$ groups (Table 3.6.10). After a period of fluctuating high fishing mortality in the mid 1990s, the F3-6 values has slightly declined and stabilized around 0.4-0.5. After a marked decline in the mid 1990s and a slight increase after the late 1990s the SSB is now fluctuating at around $140000-180000 \mathrm{t}$ (Table 3.6.9).

## Overall trends in the age structured data for the ICA model

Exploring the cohort dynamics by log catch and log survey indices gives an indication of total mortality and catchability in successive cohorts from year classes 1991-2003 (Figures 3.6.9ad). The slopes of log catches indicate a continuous decreasing trend in mortality (Figure 3.6.9a). Slopes from the three surveys; Division IIIa acoustic survey (Figure 3.6.9b), the Subdivisions 22-24 acoustic survey (Figure 3.6.9c), and the IBTS in quarter 3 in the Kattegat (Figure 3.6.9d) are more fluctuating, with a tendency in the latest years to a decreasing trend in mortality (Figures 3.6.9c-d). Although these cohorts are still based on few age-classes (3-4), all the slopes have $\mathrm{R}^{2}>0.9$. There is therefore no indication of a long term increase in total mortality based on these indices.

### 3.7 Short term projections

The assessment was used to provide a yield-per-recruit plot for WBSS herring in Division IIIa and Subdivisions 22-24 (Figure 3.7.1). The values for $\mathbf{F}_{0.1}$ and $\mathbf{F}_{\max }$ are 0.22 and 0.53 respectively, although $\mathbf{F}_{\text {max }}$ is not well defined.

Short-term predictions were carried out using MFDP v.1a software. ICA estimates of population numbers and fishing mortalities were used except for the numbers of 0 -ringers in 2006-2009, where the geometric mean of the recruitment over the period 2000-2004 was taken, and for the numbers of 1 -ringers in 2007, where the geometric mean over the period 2001-2005 was used. A shorter period compared to recent years was used in the short-term predictions. This choice reflects the decreasing trend in recruitment observed in the last years. A similar trend is observed for the North Sea herring.

Mean weights-at-age in the catch and in the stock were taken as a mean for the years 2004-2006. A status quo fishing mortality for 2007 onwards was assumed, with values rescaled to the last year estimate. Input data for catch predictions are presented in Table 3.7.1. It is worth of notice that Status quo F in 2006 is around $\mathbf{F}_{\text {max.. }}$

Short-term predictions were carried out assuming a status quo fishing mortality for 2007. The single option table is available for 2007 to 2009 (Table 3.7.2) for the following scenarios: 1) Status quo F, 2) $\mathrm{F}_{01}$ and 3) F according to a $15 \%$ catch reduction in 2008 and Status quo F in $2009\left(\mathrm{~F}_{2009}=\mathrm{F}_{2008}\right)$.

| Scenario | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: |
| 1) status quo $F$ | $\mathrm{F}_{2007}=\mathrm{F}_{2006}=0.522$ <br> Status quo F $\text { Catch }=89800 \mathrm{t}$ | $\mathrm{F}_{2008}=\mathrm{F}_{2007}=0.522$ <br> Status quo F $\text { Catch }=85500 \mathrm{t}$ | $\mathrm{F}_{2009}=\mathrm{F}_{2008}=0.522$ <br> Status quo F $\text { Catch }=86300 \mathrm{t}$ |
| 2) $\mathrm{F}_{01}$ | $\mathrm{F}_{2007}=\mathrm{F}_{2006}=0.522$ <br> Status quo F $\text { Catch }=89800 \mathrm{t}$ | $\begin{aligned} & \mathrm{F}_{2008}=0.219 \\ & \mathrm{~F}_{01} \\ & \text { Catch }=42200 \mathrm{t} \end{aligned}$ | $\begin{aligned} & \mathrm{F}_{2009}=\mathrm{F}_{2008}=0.219 \\ & \mathrm{~F}_{01} \\ & \text { Catch }=48800 \mathrm{t} \end{aligned}$ |
| 2) $\mathrm{F} \approx$ Catch reduction (-15\%) in 2008 and Status quo F in 2009 | $\mathrm{F}_{2007}=\mathrm{F}_{2006}=0.522$ <br> Status quo F $\text { Catch }=89800 \mathrm{t}$ | $\begin{aligned} & \mathrm{F}_{2008}=0.455 \\ & \mathrm{~F}_{(-15 \%)} \\ & \text { Catch }=76300 \mathrm{t} \end{aligned}$ | $\mathrm{F}_{2009}=\mathrm{F}_{2008}=0.455$ <br> Status quo F $\text { Catch }=81300 \mathrm{t}$ |

The results of the short-term predictions are given in Tables 3.7.2-3.7.5. Table 3.7.2 shows single option predictions for 2007-2009. Table 3.7.3 shows multiple options for 2008 at status quo fishing mortality in 2007. The catches for 2008 and 2009 at status quo fishing mortality were predicted to be 85500 t and 86300 t , respectively, which is an overall slight decrease in relation to the current catch level of 88900 t . The SSB is predicted to decrease to 135300 t in 2008 and to 134800 t in 2009 , which are slightly larger than the lowest values observed during the middle of the 1990's. Based on $\mathrm{F}_{0.1}$ (0.22), SSB in 2008 and 2009 are predicted to be around 139000 t and 176000 . This corresponds to landings of 42200 t in 2008 and 48800 t in 2009. Predictions based on catch reductions ( $-15 \%$ ) scenario would result
in SSB around 136200 t and 143000 t in 2008 and 2009. This would generate landings of about 76300 t and 81300 t in 2008 and 2009, respectively.

### 3.8 Precautionary and yield based reference points

The estimated SSB has not been below 111615 t since 1991. However, reference points for F and SSB have neither been defined nor proposed for this stock as

- there is no obvious stock-recruitment relationship and
- the time series is quite short with revised catch data and reliable splitting factors for only 15 years.

To find appropriate reference points simulations have been performed this year. The method proposed and introduced performs IIIa herring management scenarios using a simulation and risk based non-linear optimization procedure. It allows estimating optimum F (or effort) values to be applied and optimum catch values to be taken. As such, it holds the potential to derive precautionary reference points.

The method uses past abundance values of the last assessment year as start values to initialize the optimization process (these start values, however, may also come from direct observations such as research surveys, commercial surveys, etc.). But the method differs from conventional scenario based medium-term projection models as it is not a prediction model evaluating the influence of measures taken as part of an impact study and using past $F$ values that have been estimated from the stock history. However, in contrast to conventional scenarios the optimization tries to directly control a fishery by generating optimum F values. Along with these, optimum catches (optimum quotas or TAC values) will be generated as a result of the optimization process, given a priori specified upper F and lower SSB limits, respectively. I.e. based on this the outcome in terms of F and TAC values is expected to reflect the best management strategy under the given conditions. In case of IIIa herring it does this by maximizing the following objective function
Objective Function $=$ total herring catch

$$
\begin{align*}
& -\lambda_{1} \times \max \left(0, \mathrm{SSB}_{\text {lower limit }}-\mathrm{SSB}_{\text {estimated }}\right)  \tag{3.8.1}\\
& -\lambda_{2} \times(1-0.15) \times \max \left(0,\left(\text { catch }- \text { catch }_{\text {mean }}\right)^{2}\right)
\end{align*}
$$

The equation consists of three components in which the latter two are penalizing either an undercut of the lower SSB limit or a change of resulting catches that fall outside a symmetrical $\forall 15 \%$ interval being the current catch variation allowance suggested by the EU commission. The two $\lambda$ values are weight factors that can be used for strengthening or relaxing the penalties associated with them. Beyond this the other equations describing the underlying population dynamics are given in Gröger (2007). However, once the optimum F values are found, the catch calculated can be considered to be an optimum catch and thus may be used as TAC recommendation.

All values being input to the optimization procedure such as initial abundance values, weights, maturity observations, recruitment, etc. are prone to errors. Errors in general create uncertainty and uncertainty creates risk. The errors can be of systematic (bias) or of random nature (stochasticity). The errors address the fact that initial abundance values might have been overestimated, or that recruitment varies randomly with some variance around some function or expected value chosen. Investigating the underlying recruitment pattern of IIIa herring based on existing data for years 1991 to 2006 leads to the assumption that the IIIa herring recruitment is purely random. This allows the optimization process to be setup as being driven by a recruitment component that can be generated from a normal distribution with mean 3.94 E 9 and standard deviation 1.37E9 ( $\mathrm{p}_{\mathrm{H} 0: \mathrm{R} \sim \mathrm{N}}>0.250$ ) without an assumption of $1^{\text {st }}$ or $2^{\text {nd }}$ order autocorrelation ( $\mathrm{p}_{\mathrm{DW}}<0.0001$ ) (see Figures 3.8.1, 3.8.2, and 3.8.3). Thus, to initiate the
process of generating random recruitment values from a $\mathrm{N}[3.94 \mathrm{E} 9,1.37 \mathrm{E} 9]$ the seed values were taken in a way that the generated streams of random numbers will be annually independent of each other:

$$
\begin{equation*}
\operatorname{seed}_{2}=\operatorname{int}\left(10000 \text { x CDF }_{\text {uniform }}^{-1}\left(\text { seed }_{1}\right)\right) \tag{3.8.2}
\end{equation*}
$$

with $\mathrm{CDF}^{-1}$ uniform being the quantile function (inverse cumulative distribution function) of the uniform distribution. According to this setup of stochasticity, re-running the optimization procedure many times will result in different outcomes. Depending on the process specification, this may also lead to violations of the biological constraints as the SSB limit set may be undercut. In general, such undercuts can be interpreted as negative (hazardous, harmful) events and will happen with some frequency, in probabilistic terms with some likelihood. However, to get the optimization running, some other important points must be addressed from which specifications necessary for the optimization procedure need to be derived and setup. These issues are:

- As there are no reference points currently set for IIIa herring, what might be appropriate levels of F and SSB for specifying reference points? That includes
- a specification of a lower SSB limit not to be undercut during the optimization process
- a specification of an upper $F$ limit below which the $F$ value optimization can take place
- Is there an indication of changes in individual growth over time that may influence biomass production in time significantly?
- Is there an indication for a potential underreporting of catches (as part of an implementation error)?
- Is there an indication for potential changes (increases) in M?
- Is there an indication for potential hidden trends in F that may be induced by technological development and add to the implementation error?

To deal with these issues, we varied the assumptions relevant to the scenario and optimization based procedures by applying the following range of values:

- Candidates for SSB ref. points (lower limits): 111615 tons (Minimum), 156714 tons ( $\mathrm{P}_{50}$ ), 177609 tons (Mean)
- Candidates for F ref. points (upper limits): $0.22\left(\mathrm{~F}_{0.1}\right), 0.27,0.32,0.41$ (status quo) and 0.46 ( $\mathrm{F}_{\mathrm{Max}}$ )
- Underreporting of catch (adding to the implementation error): by $0 \%, 20 \%, 40 \%$
- Trends in mean weight: 0 and -0.00093112 (slope in mean catch weight, see Figures 3.8.4 and 3.8.5)
- Hidden trends of F (adding to the implementation error): 0 and 0.01 (increase of technological efficiency)
- Changes in M: by $0 \%$ and $20 \%$

The simulations were generally based on a planning horizon of 10 years, starting with year 2007. Since our starting point is the year 2007, we based the optimization on input data from year 2006. The relevant stock data were hence taken from the most recent stock assessment of IIIa herring as reported by the Herring Assessment Working Group (HAWG) for year 2007. All data used are age disaggregated (ages 0-8) consisting of abundance estimates $\mathrm{N}_{2006}$, weight W in kg , selectivity S , maturity observations and natural mortality M . The abundance estimates are based on final ICA estimates (Integrated Catch Analysis) derived from the 2007 HAWG assessment.

Firstly, simple but fast running scenario based simulations were performed, using 1000 runs each. Secondly, given their results, more complex optimization based simulations have been
designed and performed, using 100 runs each. The scenario based simulation runs thus served only to find initial conditions for the optimization runs. The setup and initial conditions of the optimization type simulations can be summarized as follows:

1. Optimization A. (baseline scenario with SSB lower limit set to minimum SSB and F upper limit to status quo overall F)

- $\quad$ SSB lower limit $=111615$ tons $($ minimum SSB between 1991 and 2006)
- $\quad$ F upper limit $=0.41$ (for ages $0-8$ )
- $40 \%$ underreported catch
- Average hidden trend of F of 5\% due to technological development
- M 20\% higher

2. Optimization B. (SSB lower limit set to mean SSB )

- $\quad$ SSB lower limit $=177609$ tons (mean SSB between 1991 and 2006)
- F upper limit $=0.41$ (for ages $0-8$ )
- $40 \%$ underreported catch
- Average hidden trend of F of 5\% (due to technological development)
- M 20\% higher

3. Optimization C. (SSB lower limit set to mean SSB and F upper limit decreased from status quo overall F to 0.32 )

- SSB lower limit $=177609$ tons (mean SSB between 1991 and 2006)
- $\quad$ F upper limit $=0.32$ (for ages $0-8$ )
- $40 \%$ underreported catch
- Average hidden trend of F of 5\% (due to technological development)
- M 20\% higher

4. Optimization D. (F upper limit decreased from 0.32 to 0.27 )

- SSB lower limit = 177609 tons (mean SSB between 1991 and 2006)
- F upper limit $=0.27$ (for ages $0-8$ )
- $40 \%$ underreported catch
- Average hidden trend of F of 5\% (due to technological development)
- M 20\% higher

5. Optimization E. (F upper limit decreased from 0.27 to $\mathrm{F}_{0.1}=0.22$ )

- SSB lower limit $=177609$ tons (mean SSB between 1991 and 2006)
- $\quad$ F upper limit $=0.22$ (for ages $0-8$ )
- $40 \%$ underreported catch
- Average hidden trend of F of 5\% (due to technological development)
- M $20 \%$ higher

While the first optimization run was based on the lowest SSB ever observed during period 1991 to 2006, all other runs were based on the long term average SSB estimated from years 1991 to 2006 as being the lower SSB limit allowed. The hierarchical design of the optimization runs then reflect a gradual decrease in the upper F limit (beginning with status quo overall $\mathrm{F}=0.41$ and ending with $\mathrm{F}_{0.1}=0.22$ ) below which the F values get optimized. This is to find that point below which the risk of undercutting the lower SSB limit can be ignored. The results are illustrated by Figures 3.8 .6 to 3.8.10. All five figures similarly consist of 4 panels in which

- the upper left panel always shows the trajectories of R and SSB over time (plus their minimum and maximum values (vertical lines)),
- the upper right panel the trajectories of catch (TAC) and its changes (TAC change) over time (plus a $\pm 15 \%$ interval as it was suggested by the EU commission as well as minimum and maximum values (vertical lines)),
- the lower left panel the trajectories of the optimized F values over time (plus their minimum and maximum values (vertical lines)), and
- the lower right panel the trajectories of the likelihood of undercutting the lower SSB limit, the mean loss in SSB and the combined risk over time.

The results thus indicate that an overall F value should be around $\mathrm{F}=0.27$ as the expected risk of SSB falling below the long term mean of SSB is negligibly small (around 1.5 on average,
see Figure 3.8.9). This means a probability of occurrence of $1.1 \%$ and an average (future) loss of SSB of 137 tons per year. The loss addresses the fact that lost SSB will be missing in the next year's SSB budget which does affect the regeneration potential but also the future catches only marginally.

### 3.9 Quality of the Assessment

The assessment in 2007 is an update of last year's assessment. Therefore, the assessment has not been explored beyond examining the standard diagnostics.

Three data series (surveys) are used in addition to the catch numbers at age. None of these surveys cover the whole distribution area of the stock, but each of them covers areas where it is likely that certain ages are well represented at survey time. The acoustic survey in Division IIIa+IVaE covers fish age 2 and older while the two others largely cover the younger part of the population. Hence, these surveys can be regarded as complementary. All surveys are noisy. The acoustic survey in Division IIIa +IVaE indicates a higher mortality than the others, but its contribution to the total sum of squares does not have a distinct minimum (Figure 3.6.5). The selection pattern is smooth and no age (1+) or year effects (2001-2005) in model residuals are large (Figure 3.6.7).

Altogether, the current procedure for assessing the stock has given consistent results with respect to fishing mortality, spawning biomass and recruitment for several years (Figure 3.9.1).

The retrospective errors are small, except in the recruitment and even these are unbiased (Figure 3.9.2). Apparently, the strength of a year class is not firmly estimated before the year class has been followed for 2-3 years. The selection at age in the fishery changes in retrospective runs. This probably reflects a stronger exploitation of younger herring in earlier years, which in the present assessment is reflected in the VPA part. The selection at age in this year's assessment is similar to that in two last year's assessment (Figure 3.9.3), while the tendency to a decrease in exploitation for the younger ages is confirmed and the catch residuals are relatively small. Hence, the separable assumption does not seem to be generally violated.

Single fleet ICA runs show that SSB estimates from the final run for 2007 are lying between those obtained using IBTS Q3 and Q1, with Acoustic tuning fleets being closer to the SSB values of the final run (Figure 3.9.4) when compared to IBTS.

For prediction purposes, better indicators of recruitment would be useful. At present, geometric mean recruitment has to be assumed for age 0 in the intermediate year and for later years. HAWG suggests to investigate procedures that give a better predictive power of the recruitment by reducing the impact of outliers and to analyse within survey variances. The predictions are made for the Western Baltic Spring Spawning (WBSS) stock, while management is by areas. In Division IIIa, the fishery exploits both WBSS and North Sea autumn spawning herring. The Working Group has attempted to outline the consequences for both stocks in fishery in Division IIIa (Section 3.10). This requires insight to both how the catches of WBSS are distributed by areas, and the proportions of the catches in Division IIIa from each stock. Both these properties change over time, and are influenced both by managers' decisions and the abundance of the respective stocks in the area. So far, the only basis has been historical data of catches in biomass by area and species (cfr. Table 2.1.6). A better basis could be achieved by considering catches at age by different fleets, and investigations of how management decisions influence the fishery. Further a deeper understanding of relationships between stock characteristics and major migration patterns would help predictions of the seasonal stock composition in the mixed areas. These efforts
require inter-sectional work; an attempt to resolve parts of the problem has been addressed through the IAMHERSKA project (see section 1.4.8).

Compared to last year's assessment, the change in the estimate is $+15 \%$ and $+21 \%$ for the fishing mortalities in 2004 and 2005 respectively; and $-6 \%$ and $-11 \%$ for the SSB in 2004 and 2005 respectively. The text table below gives an overview of the assumptions made in the 2006 and 2007 assessments and a comparison of the main results with 2004 and 2005 as baselines.

| CATEGORY | PARAMETER | AsSESSMENT IN 2006 | Assessment in 2007 | $\begin{gathered} \text { DIFF. 07-06 } \\ \text { (+/-) \% } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: |
| ICA input | No. of years for separable constraints | 5 | 5 | No |
|  | Reference age for separable constraint | 4 | 4 | No |
|  | Selection to be fixed on last age | 1 | 1 | No |
|  | Weighting factor to all indices | 1 | 1 | No |
|  | Catch downweighted to 0.1 for 0 ringer | Yes | Yes | No |
|  | Tuning data | Acoustic Surv. Div. IIIa 2-8+ ringers | Acoustic Surv. Div. IIIa 2-8+ ringers | No |
|  |  | ```Acoustic Surv. SDs 22- 24 0-5 ringers``` | ```Acoustic Surv. SDs 22- 24 0-5 ringers``` | No |
|  |  | IBTS Surv. Quarter 3 1-5 ringers | IBTS Surv. Quarter 3 1-5 ringers | No |
| ICA results | $\begin{gathered} \text { SSB } 2004 \\ \text { F(3-6) } 2004 \end{gathered}$ | 168,700 t | 158,200t | -6\% |
|  |  | 0.386 | 0.444 | +15\% |
|  | $\begin{gathered} \text { SSB } 2005 \\ \mathrm{~F}(3-6) 2005 \end{gathered}$ | 164,600 t | 155,300 t | -6\% |
|  |  | 0.408 | 0.494 | +21\% |

### 3.10 Management Considerations

## Catch options for mixed stocks in Division IIIa based on short term predictions for WBSS

There is an indication of a declining recruitment in recent years in the WBSS herring stock. The present state of a declining NSAS stock with poor recruitment in the last 5 years strongly suggest that advice given for the WBSS stock will not conflict with the present comanagement of the two stocks in the mixed areas of Division IIIa and Division IVaE.

It should however also be noted that the scope for exploitation is not only dependent on the overall population dynamics of the two stocks. Management also has to consider age-class specific stock composition in the mixing zones brought about by unpredictable changes in distribution pattern triggered by environmental as well as population biological and behavioural cues.

The current fleet definitions are:

## North Sea

Fleet A: Directed herring fisheries with purse seiners and trawlers. By-catches in industrial fisheries by Norway are included.
Fleet B: Herring taken as by-catch under EU regulations.

## Division IIIa

Fleet C: Directed herring fisheries with purse seiners and trawlers
Fleet D: By-catches of herring caught in the small-mesh fisheries

Subdivision 22-24
Fleet F: All herring fisheries in Subdivisions 22-24

## Quotas in Division IIIa

The quota for the C-fleet and the by-catch quota for the D-fleet are set for both stocks together. Therefore the implication of the quotas for the outtake of WBSS has to be considered. Furthermore the implication for the outtake of NSAS has to be taken into account when setting fleet wise quotas for that stock (see section 2.7).

For 2006 the agreed TAC for the directed fishery in Division IIIa (C-fleet) was 81600 t . The TAC was divided into quotas, 500 t for the Faeroes, 70217 t for the EU of which all had to be taken in Division IIIa, and 10883 t for Norway of which $50 \%$ could be taken in the North Sea. A by-catch quota for Division IIIa herring in the small meshed fishery (fleet-D) was set at 22528 t .

For 2007 the agreed TAC for the directed fishery in Division IIIa (C-fleet) is 69360 t . The TAC is divided into quotas, 500 t for the Faeroes and 9251 t for Norway of which $40 \%$ can be taken in the North Sea. For the EU a total quota of 59609 t is agreed of which all has to be taken in Division IIIa; this is divided between Denmark 28907 t , Germany 463 t and Sweden 30239 t . A by-catch quota for Division IIIa herring in the small meshed fishery (fleet-D) is set at 15396 t and divided between Denmark 13160 t , Germany 117 t and Sweden 2119 t .

It must also be noted that a slightly variable and relatively small amount (around 8000 t ) of WBSS herring is taken in the fishery in Subarea IV (see Section 2.2.2 and Figure 2.2.2 for information about WBSS taken in Divisions IVa and IVb East). This component is accounted for in both the assessments on NSAS and WBSS. Adding to this there is misreporting by areas. In recent years, HAWG has calculated a substantial part of the catch reported as taken in Division IIIa in fleet C actually has been taken in Subarea IV. These catches have been allocated to the North Sea stock and accounted under the A-fleet. Regulations allowing quota transfers from Division IIIa to the North Sea were introduced with the incentive to decrease misreporting for the Norwegian part of the fishery. However, working group estimates suggest that out of the official landings for human consumption in the Skagerrak, $46 \%, 58 \%, 46 \%$ and $36 \%$ are misreported in 2003, 2004, 2005 and 2006 respectively. These figures are probably underestimating the problem since only a subset of countries supply this information to the HAWG. Misreported catches are moved to the appropriate stock for the assessment.

## TAC in Subdivisions 22-24

For 2007 the agreed TAC for the herring fishery in Subdivisions 22-24 (Fleet F) is 49500 t . The TAC is divided into quotas, 6939 t for Denmark, 27311 t for Germany, 3 t for Finland, 6441 t for Poland and 8806 t for Sweden.

## ICES catch predictions versus management TAC

ICES gives advice on catch options for the entire distribution of the two herring stocks separately, whereas herring is managed by areas cross sectioning the geographical distribution of the stocks (see the following text diagram).


## Data used for catch options in 2008

There is no firm basis for predicting the fraction of NSAS in the catches by the C- and Dfleets. The proportions of the two stocks as well as the distribution pattern of the fishery in the Eastern North Sea and the Division IIIa is dynamically changing year by year. This is probably influenced by year-class strength of the two stocks and their relative geographical distributions as well as fleet behaviour reacting on herring availability and management decisions.

Recent years' shares of the WBSS catches in IIIa and other areas is used to translate the total recommended TAC for WBSS into outtake of WBSS in Division IIIa and Subdivisions 22-24. The mix of the two stocks in the Division IIIa catches is used to derive the outtake of NSAS and total catches in Division IIIa. Predicted catches of WBSS and NSAS by fleet in IIIa is based on recent patterns of 1) ratio of WBSS catches taken by each fleet and 2) proportion of the two stocks in catches of the different fleets.

The catch option for 2008 is based on the share by fleet and stock composition in catches given as a mean for the years 2004-2006. The ratio by fleet and stock composition is given in the following text table A and B, respectively:

Text table A showing the 2004-2006 average share of the total catch in $t$ of WBSS by each fleet.

| WBSS | Fleet C (IIIa) | Fleet D (IIIa) | Fleet F (SD22-24) <br> + Fleet $\boldsymbol{A}(\boldsymbol{I V})^{*}$ | Total |
| :--- | :---: | :---: | :---: | :---: |
| Mean $(2004-2006)$ <br> catch in t | 26,500 | 7,400 | 54,800 | 88,700 |
| Mean $(2004-2006)$ <br> share in $\%$ | $30 \%$ | $8 \%$ | $62 \%$ | $100 \%$ |

*A constant catch of 8400 t of WBSS caught in Subarea IV is accounted for in the calculations.
Text table B showing the 2004-2006 average proportion of WBSS in catches by fleet (the split).

| WBSS | Fleet C | Fleet D | Fleet F (SD22-24) <br> + Fleet $\boldsymbol{A}$ (IV) |
| :--- | :---: | :---: | :---: |
| Mean (2004-2006) <br> proportion | 0.62 | 0.49 | 1.00 |

## Exploring a range of total WBSS catches

The settings of $F$ in the stock short term projections considered the indication of declining recruitment in the WBSS stock as well as the present level of a declining NSAS stock with very low recruitment in recent years. Catch options with focus on incremental change towards management for a maximum sustainable yield were explored for the two stocks in Division IIIa at total catches set for the WBSS stock.

The projected stock composition is assumed to equal the 2004-2006 average of the NSAS and WBSS in each of the C and D fleets (in Division IIIa). Further the 2004-2006 average catch of 8400 t of WBSS is assumed taken in Subarea IV.

The stock assessment indicates a recent increase in SSB to 185000 t in 2006 likely driven by the quite large 2003 year-class coming through. Short-term projections calculate that the assumed catch in 2007 (in total 90000 t with status quo fishing mortality) will lead to a decrease in SSB in 2007 to about 154000 t (Table 3.7.2). Catch options for 2008 and 2009 with $\mathrm{F}_{\mathrm{sq}}$ will further decrease SSB in 2008 and in 2009 (Table 3.7.2), whereas a fishing mortality based on a $15 \%$ catch reduction ( $\mathrm{F}_{\mathrm{C}-15 \%}$ ) will lead to a decrease in SSB in 2008 and a
slight increase in 2009 (Table 3.7.4). The setting of the more restrictive $\mathrm{F}_{01}$ will lead to a decrease in SSB in 2008 and an increase in 2009 (Table 3.7.3).

The text table below gives catch options based on the F levels above and a series of other scenarios derived from the HAWG2007 short-term projections for the WBSS in Division IIIa, in SDs 22-24 and in Subarea IV.

In the text table below the options in bold corresponds to the three F - scenarios: Option $1 \approx$ $\mathrm{F}_{0.1}$, option $5 \approx \mathrm{~F}_{\mathrm{C}-15 \%}$, and option $7 \approx \mathrm{~F}_{\mathrm{sq}}$, further a number of other options between 40200 t and 88700 t are given (values are rounded to the nearest 100 t ).

| MANAGEMENT CONSIDERATIONS FOR DIVISION IIIA + SD 22-24 BASED ON SHORT TERM PREDICTIONS (HAWG 2007) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catch option for the WBSS herring stock |  | WBSS herring |  |  |  | NSAS herring |  | Total catches of both stocks in Division IIIa and Sub-division 22-24 |  |  |
| Option | Total catches of WBSS herring ${ }^{*}$ | FleetA* | Fleet C | Fleet D | Fleet F | Fleet C | FleetD | Fleet <br> C | Fleet <br> D | Fleet <br> F |
| $1\left(\mathrm{~F}_{0.1}\right)$ | 40,200 | 8,400 | 12,600 | 3,500 | 15,800 | 7,600 | 3,700 | 20,200 | 7,200 | 15,800 |
| 2 | 50,000 | 8,400 | 15,700 | 4,400 | 21,500 | 9,400 | 4,500 | 25,100 | 8,900 | 21,600 |
| 3 | 60,000 | 8,400 | 18,800 | 5,200 | 27,600 | 11,300 | 5,500 | 30,100 | 10,700 | 27,600 |
| 4 | 70,000 | 8,400 | 21,900 | 6,100 | 33,600 | 13,200 | 6,400 | 35,100 | 12,500 | 33,600 |
| $5\left(\mathrm{~F}_{\mathrm{C}-15 \%}\right)$ | 76,300 | 8,400 | 23,900 | 6,600 | 37,400 | 14,400 | 6,900 | 38,300 | 13,600 | 37,400 |
| 6 | 85,000 | 8,400 | 26,600 | 7,400 | 42,600 | 16,000 | 7,700 | 42,600 | 15,100 | 42,600 |
| $7\left(\mathrm{~F}_{\mathrm{sq}}\right)$ | 88,700 | 8,400 | 27,800 | 7,700 | 44,800 | 16,700 | 8,100 | 44,500 | 15,800 | 44,800 |
| *A catch of 8400 t of WBSS herring taken in the Eastern North Sea is assumed. |  |  |  |  |  |  |  |  |  |  |

The short term projection with recent catch levels show a decline in SSB which indicates that fishing mortality should be reduced. Catches based on $\mathrm{F}_{01}$ quickly re-establish the SSB to above average values at the cost of high reductions in yield. Applying an incremental approach towards $\mathrm{F}_{0.1}$ (as a proxy for maximum sustainable yield) may be achieved on a longer term basis by successive $15 \%$ catch reductions provided recruitment levels are not further reduced. However a catch reduction of $15 \%$ in 2008 followed by a status quo F in 2009 appears to be a rather slow progress in that direction.

For a TAC on catch of NSAS and total catch by the fleets in Division IIIa to be compatible with the advice for WBSS, the numbers derived as above, based on the largest advisable catch of WBSS, are upper bounds on the advisable catches of NSAS by the C- and D- fleets. Thus the resulting catch options were also used as constraints for short term predictions for the NSAS herring (section 2.7).

Table 3.1.1 WESTERN BALTIC HERRING.
Total landings in 1986-2006 in thousands of tonnes.
(Data provided by Working Group members 2006).

| Year | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Skage rrak |  |  |  |  |  |  |  |  |  |  |
| Denmark | 94.0 | 105.0 | 144.4 | 47.4 | 62.3 | 58.7 | 64.7 | 87.8 | 44.9 | 43.7 |
| Faroe Islands | 0.5 |  |  |  |  |  |  |  |  |  |
| Germany |  |  |  |  |  |  |  |  |  |  |
| Norway | 1.6 | 1.2 | 5.7 | 1.6 | 5.6 | 8.1 | 13.9 | 24.2 | 17.7 | 16.7 |
| Sweden | 43.0 | 51.2 | 57.2 | 47.9 | 56.5 | 54.7 | 88.0 | 56.4 | 66.4 | 48.5 |
| Total | 139.1 | 157.4 | 207.3 | 96.9 | 124.4 | 121.5 | 166.6 | 168.4 | 129.0 | 108.9 |
| Kattegat |  |  |  |  |  |  |  |  |  |  |
| Denmark | 37.4 | 46.6 | 76.2 | 57.1 | 32.2 | 29.7 | 33.5 | 28.7 | 23.6 | 16.9 |
| Sweden | 35.9 | 29.8 | 49.7 | 37.9 | 45.2 | 36.7 | 26.4 | 16.7 | 15.4 | 30.8 |
| Total | 73.3 | 76.4 | 125.9 | 95.0 | 77.4 | 66.4 | 59.9 | 45.4 | 39.0 | 47.7 |
| Sub. Div. 22+24 |  |  |  |  |  |  |  |  |  |  |
| Denmark | 14.0 | 32.5 | 33.1 | 21.7 | 13.6 | 25.2 | 26.9 | 38.0 | 39.5 | 36.8 |
| Germany | 60.0 | 53.1 | 54.7 | 56.4 | 45.5 | 15.8 | 15.6 | 11.1 | 11.4 | 13.4 |
| Poland | 12.3 | 8.0 | 6.6 | 8.5 | 9.7 | 5.6 | 15.5 | 11.8 | 6.3 | 7.3 |
| Sweden | 5.9 | 7.8 | 4.6 | 6.3 | 8.1 | 19.3 | 22.3 | 16.2 | 7.4 | 15.8 |
| Total | 92.2 | 101.4 | 99.0 | 92.9 | 76.9 | 65.9 | 80.3 | 77.1 | 64.6 | 73.3 |
| Sub. Div. 23 |  |  |  |  |  |  |  |  |  |  |
| Denmark | 1.5 | 0.8 | 0.1 | 1.5 | 1.1 | 1.7 | 2.9 | 3.3 | 1.5 | 0.9 |
| Sweden | 1.4 | 0.2 | 0.1 | 0.1 | 0.1 | 2.3 | 1.7 | 0.7 | 0.3 | 0.2 |
| Total | 2.9 | 1.0 | 0.2 | 1.6 | 1.2 | 4.0 | 4.6 | 4.0 | 1.8 | 1.1 |
| Grand Total | 307.5 | 336.2 | 432.4 | 286.4 | 279.9 | 257.8 | 311.4 | 294.9 | 234.4 | 231.0 |


| Year | 1996 | 1997 | $1998{ }^{2}$ | $1999{ }^{2}$ | 2000 | $2001{ }^{5}$ | $2002{ }^{4}$ | 2003 | 2004 | 2005 | $2006{ }^{1,3}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Skagerrak |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 28.7 | 14.3 | 10.3 | 10.1 | 16.0 | 16.2 | 26.0 | 15.5 | 11.8 | 14.8 | 5.2 |
| Faroe Islands |  |  |  |  |  |  |  |  |  | 0.4 |  |
| Germany |  |  |  |  |  |  |  | 0.7 | 0.5 | 0.8 | 0.6 |
| Norway | 9.4 | 8.8 | 8.0 | 7.4 | 9.7 |  |  |  |  |  |  |
| Sweden | 32.7 | 32.9 | 46.9 | 36.4 | 45.8 | 30.8 | 26.4 | 25.8 | 21.8 | 32.5 | 26.0 |
| Total | 70.8 | 56.0 | 65.2 | 53.9 | 71.5 | 47.0 | 52.3 | 42.0 | 34.1 | 48.5 | 31.8 |
| Kattegat |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 17.2 | 8.8 | 23.7 | 17.9 | 18.9 | 18.8 | 18.6 | 16.0 | 7.6 | 11.1 | 8.6 |
| Sweden | 27.0 | 18.0 | 29.9 | 14.6 | 17.3 | 16.2 | 7.2 | 10.2 | 9.6 | 10.0 | 10.8 |
| Total | 44.2 | 26.8 | 53.6 | 32.5 | 36.2 | 35.0 | 25.9 | 26.2 | 17.2 | 21.1 | 19.4 |

Sub. Div. 22+24

|  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Denmark | 34.4 | 30.5 | 30.1 | 32.5 | 32.6 | 28.3 | 13.1 | 6.1 | 7.3 | 5.3 | 1.4 |
| Germany | 7.3 | 12.8 | 9.0 | 9.8 | 9.3 | $\mathbf{1 1 . 4}$ | 22.4 | 18.8 | 18.5 | 21.0 | 22.9 |
| Poland | 6.0 | 6.9 | 6.5 | 5.3 | 6.6 | 9.3 | - | 4.4 | 5.5 | 6.3 | 5.5 |
| Sweden | 9.0 | 14.5 | 4.3 | 2.6 | 4.8 | 13.9 | 10.7 | 9.4 | 9.9 | 9.2 | 9.6 |
| Total | 56.7 | 64.7 | 49.9 | 50.2 | 53.3 | 62.9 | 46.2 | 38.7 | 41.2 | 41.8 | 39.4 |

Sub. Div. 23

| Denmark | 0.7 | 2.2 | 0.4 | 0.5 | 0.9 | 0.6 | 4.6 | 2.3 | 0.1 | 1.8 | 1.8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sweden | 0.3 | 0.1 | 0.3 | 0.1 | 0.1 | 0.2 | - | 0.2 | 0.3 | 0.4 | 0.7 |
| Total | 1.0 | 2.3 | 0.7 | 0.6 | 1.0 | 0.8 | 4.6 | 2.6 | 0.4 | 2.2 | 2.5 |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| Grand Total | 172.7 | 149.8 | 169.4 | 137.2 | 162.0 | 145.7 | 128.9 | 109.5 | 92.8 | 113.6 | 93.0 |

${ }^{1}$ Preliminary data.
2 Revised data for 1998 and 1999
Bold= German revised data for 2001
${ }^{3} 2000$ tonnes of Danish landings are missing, see text section 3.1.2
${ }^{4}$ The Danish national management regime for herring and sprat fishery in Subdivision 22 was changed in 2002
${ }^{5}$ The total landings in Skagerrak have been updated for 1995-2001 due to Norwegian misreportings into Skagerrak

Table 3.1.2
WESTERN BALTIC HERRING.
Landings (SOP) in 2001-2006 by fleet and quarter (1000 t).

| Year | Quarter | Div. Illa |  | SD 22-24 | Div. Illa + SD 22-24 |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Fleet C | Fleet D | Fleet F | Total |
| 2001 | 1 | 19.6 | 3.8 | 20.8 | 44.2 |
|  | 2 | 11.1 | 1.9 | 20.7 | 33.7 |
|  | 3 | 24.7 | 7.9 | 7.5 | 40.1 |
|  | 4 | 11.1 | 1.7 | 14.8 | 27.6 |
|  | Total | 66.5 | 15.3 | 63.8 | 145.6 |
| 2002 | 1 | 11.4 | 6.2 | 19.6 | 37.2 |
|  | 2 | 6.3 | 2.1 | 18.3 | 26.7 |
|  | 3 | 23.2 | 7 | 1.5 | 31.7 |
|  | 4 | 14.2 | 2.5 | 13.3 | 30.0 |
|  | Total | 55.1 | 17.8 | 52.7 | 125.6 |
| 2003 | 1 | 10.9 | 7 | 20.3 | 38.2 |
|  | 2 | 7.9 | 1.3 | 12.9 | 22.1 |
|  | 3 | 21.9 | 0.9 | 1.5 | 24.3 |
|  | 4 | 15 | 3.3 | 5.6 | 23.9 |
|  | Total | 55.7 | 12.5 | 40.3 | 108.5 |
| 2004 | 1 | 13.5 | 2.8 | 20.4 | 36.7 |
|  | 2 | 2.8 | 3.3 | 10.4 | 16.5 |
|  | 3 | 8.2 | 10.8 | 2.4 | 21.4 |
|  | 4 | 5.9 | 5.0 | 8.6 | 19.4 |
|  | Total | 30.3 | 22.0 | 41.7 | 93.9 |
| 2005 | 1 | 16.6 | 6.1 | 20.4 | 43.1 |
|  | 2 | 3.4 | 1.9 | 15.6 | 20.9 |
|  | 3 | 23.4 | 3.4 | 1.9 | 28.7 |
|  | 4 | 12.0 | 2.6 | 5.8 | 20.5 |
|  | Total | 55.4 | 14.1 | 43.7 | 113.3 |
| 2006 | 1 | 15.3 | 5.9 | 15.1 | 36.2 |
|  | 2 | 2.6 | 0.1 | 17.2 | 19.9 |
|  | 3 | 15.7 | 0.8 | 3.0 | 19.5 |
|  | 4 | 8.3 | 2.4 | 6.5 | 17.3 |
|  | Total | 41.9 | 9.3 | 41.9 | 93.0 |

Table 3.2.1 WESTERN BALTIC HERRING. Skagerrak. Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.

| Quarter | Division: |  | Skagerrak | Year: | 2006 | Country: | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 1.95 | 31 | 11.29 | 17 | 13.24 | 19 |
|  | 2 | 24.37 | 77 | 14.25 | 74 | 38.62 | 76 |
|  | 3 | 13.87 | 109 | 11.35 | 109 | 25.22 | 109 |
|  | 4 | 3.50 | 131 | 2.87 | 131 | 6.37 | 131 |
|  | 5 | 5.79 | 187 | 4.90 | 187 | 10.69 | 187 |
|  | 6 | 0.69 | 195 | 0.58 | 195 | 1.28 | 195 |
|  | 7 | 0.25 | 220 | 0.21 | 220 | 0.46 | 220 |
|  | 8+ | 0.19 | 209 | 0.16 | 209 | 0.35 | 209 |
|  | Total | 50.61 |  | 45.62 |  | 96.23 |  |
|  | SOP |  | 5,218 |  | 3,964 |  | 9,182 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 2.81 | 41 | 0.19 | 41 | 3.00 | 41 |
|  | 2 | 11.87 | 75 | 0.81 | 75 | 12.68 | 75 |
|  | 3 | 2.95 | 105 | 0.20 | 105 | 3.15 | 105 |
|  | 4 | 0.77 | 135 | 0.05 | 135 | 0.82 | 135 |
|  | 5 | 1.22 | 170 | 0.24 | 118 | 1.46 | 162 |
|  | 6 | 0.03 | 179 | 0.00 | 179 | 0.03 | 179 |
|  | 7 | 0.11 | 179 | 0.01 | 179 | 0.12 | 179 |
|  | 8+ | 0.09 | 184 | 0.01 | 184 | 0.09 | 184 |
|  | Total | 19.84 |  | 1.51 |  | 21.34 |  |
|  | SOP |  | 1,669 |  | 128 |  | 1,796 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.39 | 38 | 4.73 | 16 | 5.12 | 18 |
|  | 1 | 36.56 | 83 | 2.87 | 47 | 39.43 | 81 |
|  | 2 | 36.03 | 103 | 0.40 | 99 | 36.43 | 103 |
|  | 3 | 24.54 | 131 | 0.18 | 129 | 24.72 | 131 |
|  | 4 | 11.86 | 148 | 0.08 | 147 | 11.93 | 148 |
|  | 5 | 5.81 | 162 | 0.12 | 177 | 5.93 | 163 |
|  | 6 | 2.74 | 175 | 0.02 | 167 | 2.76 | 175 |
|  | 7 | 2.62 | 188 | 0.02 | 187 | 2.64 | 188 |
|  | 8+ | 0.98 | 204 | 0.01 | 219 | 0.98 | 204 |
|  | Total | 121.52 |  | 8.43 |  | 129.95 |  |
|  | SOP |  | 13,842 |  | 317 |  | 14,158 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 4.25 | 15 | 1.40 | 15 | 5.64 | 15 |
|  | 1 | 28.85 | 73 | 9.49 | 73 | 38.34 | 73 |
|  | 2 | 5.82 | 104 | 1.91 | 104 | 7.73 | 104 |
|  | 3 | 6.13 | 140 | 2.02 | 140 | 8.15 | 140 |
|  | 4 | 3.38 | 155 | 1.11 | 155 | 4.49 | 155 |
|  | 5 | 2.99 | 192 | 0.98 | 192 | 3.97 | 192 |
|  | 6 | 0.31 | 216 | 0.10 | 216 | 0.42 | 216 |
|  | 7 | 0.63 | 232 | 0.21 | 232 | 0.84 | 232 |
|  | 8+ | 0.24 | 207 | 0.08 | 207 | 0.31 | 207 |
|  | Total | 52.60 |  | 17.29 |  | 69.89 |  |
|  | SOP |  | 4,976 |  | 1,636 |  | 6,612 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 4.63 | 17 | 6.13 | 16 | 10.76 | 16 |
|  | 1 | 70.17 | 76 | 23.84 | 43 | 94.01 | 67 |
|  | 2 | 78.10 | 91 | 17.37 | 78 | 95.46 | 88 |
|  | 3 | 47.49 | 124 | 13.74 | 114 | 61.23 | 122 |
|  | 4 | 19.50 | 146 | 4.11 | 138 | 23.62 | 144 |
|  | 5 | 15.80 | 177 | 6.25 | 185 | 22.05 | 180 |
|  | 6 | 3.78 | 182 | 0.71 | 197 | 4.48 | 185 |
|  | 7 | 3.61 | 198 | 0.45 | 223 | 4.06 | 200 |
|  | 8+ | 1.48 | 204 | 0.25 | 208 | 1.73 | 204 |
|  | Total | 244.57 |  | 72.85 |  | 317.42 |  |
|  | SOP |  | 25,704 |  | 6,045 |  | 31,749 |

Table 3.2.2 WESTERN BALTIC HERRING. Kattegat. Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.

| Quarter | Division: |  |  | Year: | 2006 | Country: | ALL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Kattegat |  |  |  |  |
|  | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $1$ | 1 | 15.27 | 31 | 67.22 | 17 | 82.49 | 20 |
|  | 2 | 34.81 | 68 | 6.46 | 49 | 41.28 | 65 |
|  | 3 | 37.73 | 97 | 2.54 | 102 | 40.27 | 97 |
|  | 4 | 9.17 | 118 | 0.39 | 122 | 9.55 | 118 |
|  | 5 | 11.81 | 174 | 0.63 | 170 | 12.45 | 174 |
|  | 6 | 1.26 | 183 | 0.06 | 189 | 1.32 | 183 |
|  | 7 | 0.90 | 171 | 0.02 | 147 | 0.92 | 171 |
|  | 8+ | 0.37 | 171 | 0.02 | 167 | 0.39 | 171 |
|  | Total | 111.32 |  | 77.34 |  | 188.66 |  |
|  | SOP |  | 10,081 |  | 1,892 |  | 11,973 |
| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $2$ | 1 | 6.82 | 48 | 0.42 | 17 | 7.24 | 47 |
|  | 2 | 5.23 | 77 | 0.02 | 35 | 5.25 | 76 |
|  | 3 | 1.06 | 102 | 0.00 | 144 | 1.06 | 102 |
|  | 4 | 0.37 | 99 |  |  | 0.37 | 99 |
|  | 5 | 0.12 | 144 |  |  | 0.12 | 144 |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | 8+ | 0.06 | 171 |  |  | 0.06 | 171 |
|  | Total | 13.66 |  | 0.44 |  | 14.10 |  |
|  | SOP |  | 903 |  | 8 |  | 911 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $3$ | 0 |  |  | 22.12 | 12 | 22.12 | 12 |
|  | 1 | 7.69 | 66 | 5.67 | 42 | 13.36 | 56 |
|  | 2 | 7.59 | 78 | 0.15 | 77 | 7.75 | 78 |
|  | 3 | 4.89 | 93 | 0.06 | 99 | 4.95 | 94 |
|  | 4 | 1.81 | 107 | 0.02 | 112 | 1.83 | 107 |
|  | 5 | 0.42 | 120 | 0.01 | 133 | 0.43 | 120 |
|  | 6 | 0.13 | 163 | 0.00 | 163 | 0.13 | 163 |
|  | 7 | 0.02 | 153 | 0.00 | 153 | 0.02 | 153 |
|  | 8+ |  |  |  |  |  |  |
|  | Total | 22.55 |  | 28.03 |  | 50.58 |  |
|  | SOP |  | 1,828 |  | 531 |  | 2,359 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $4$ | 0 | 3.50 | 29 | 2.36 | 15 | 5.86 | 23 |
|  | 1 | 35.12 | 58 | 7.44 | 61 | 42.56 | 59 |
|  | 2 | 7.33 | 86 | 1.66 | 89 | 8.99 | 86 |
|  | 3 | 2.72 | 122 | 0.67 | 127 | 3.39 | 123 |
|  | 4 | 0.91 | 145 | 0.24 | 147 | 1.15 | 145 |
|  | 5 | 0.26 | 172 | 0.07 | 172 | 0.33 | 172 |
|  | 6 | 0.10 | 191 | 0.03 | 191 | 0.13 | 191 |
|  | 7 | 0.15 | 231 | 0.04 | 231 | 0.20 | 231 |
|  | 8+ | 0.05 | 215 | 0.01 | 215 | 0.07 | 215 |
|  | Total | 50.14 |  | 12.55 |  | 62.68 |  |
|  | SOP |  | 3,352 |  | 788 |  | 4,139 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $T$ | 0 | 3.50 | 29 | 24.48 | 12 | 27.98 | 14 |
|  | 1 | 64.90 | 52 | 80.75 | 23 | 145.65 | 36 |
| 0 | 2 | 54.96 | 73 | 8.31 | 58 | 63.27 | 71 |
|  | 3 | 46.39 | 98 | 3.28 | 107 | 49.67 | 99 |
| $t$ | 4 | 12.25 | 117 | 0.65 | 131 | 12.90 | 118 |
|  | 5 | 12.61 | 172 | 0.71 | 170 | 13.32 | 172 |
| a | 6 | 1.49 | 182 | 0.09 | 188 | 1.58 | 182 |
|  | 7 | 1.07 | 179 | 0.06 | 204 | 1.13 | 181 |
|  | 8+ | 0.48 | 176 | 0.03 | 187 | 0.51 | 176 |
|  | Total | 197.66 |  | 118.37 |  | 316.02 |  |
|  | SOP |  | 16,164 |  | 3,219 |  | 19,383 |

Table 3.2.3 WESTERN BALTIC HERRING. Division IIIa Landings in numbers (mill.), mean weight (g.) and SOP (t) by age and quarter.

| Quarter | Division: |  |  | Year: |  | 2006 | Country | $: \quad$ ALL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | W-rings | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 10.39 | 14 | 0.20 | 16 | 4.44 | 18 | 15.04 | 15 |
|  | 2 | 11.76 | 47 | 1.02 | 43 | 14.37 | 44 | 27.15 | 45 |
|  | 3 | 10.67 | 84 | 2.12 | 80 | 33.83 | 89 | 46.63 | 87 |
|  | 4 | 6.79 | 119 | 1.64 | 104 | 21.97 | 115 | 30.40 | 115 |
|  | 5 | 3.11 | 147 | 0.64 | 136 | 9.75 | 146 | 13.50 | 146 |
|  | 6 | 1.84 | 159 | 0.30 | 156 | 6.97 | 169 | 9.11 | 167 |
|  | 7 | 1.02 | 163 | 0.34 | 170 | 7.35 | 190 | 8.71 | 186 |
|  | 8+ | 0.60 | 195 | 0.22 | 176 | 4.17 | 187 | 4.99 | 187 |
|  | Total | 46.18 |  | 6.50 |  | 102.85 |  | 155.53 |  |
|  | SOP |  | 3,436 |  | 621 |  | 11,004 |  | 15,061 |
| Quarter |  | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 0.35 | 16 | 0.00 | 19 | 4.52 | 22 | 4.88 | 22 |
|  | 2 | 0.82 | 42 | 0.03 | 47 | 13.83 | 48 | 14.68 | 48 |
|  | 3 | 1.26 | 77 | 0.05 | 75 | 42.47 | 78 | 43.78 | 78 |
|  | 4 | 1.21 | 114 | 0.04 | 101 | 43.66 | 106 | 44.91 | 106 |
|  | 5 | 0.78 | 146 | 0.01 | 115 | 18.17 | 140 | 18.96 | 140 |
|  | 6 | 0.59 | 166 | 0.01 | 141 | 15.04 | 160 | 15.64 | 160 |
|  | 7 | 0.46 | 178 | 0.01 | 167 | 11.46 | 177 | 11.93 | 177 |
|  | 8+ | 0.22 | 201 | 0.00 | 171 | 5.33 | 178 | 5.56 | 179 |
|  | Total | 5.70 |  | 0.15 |  | 154.49 |  | 160.34 |  |
|  | SOP |  | 613 |  | 13 |  | 16,615 |  | 17,240 |
| Quarter |  | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.00 | 22 |  |  | 0.00 | 22 | 0.00 | 22 |
|  | 1 | 0.00 | 51 | 2.66 | 44 | 8.12 | 41 | 10.79 | 42 |
|  | 2 | 0.00 | 79 | 1.68 | 66 | 8.51 | 59 | 10.20 | 60 |
|  | 3 | 0.01 | 112 | 2.02 | 73 | 7.20 | 62 | 9.22 | 65 |
|  | 4 | 0.00 | 130 | 1.61 | 65 | 11.20 | 52 | 12.81 | 54 |
|  | 5 | 0.00 | 143 | 0.47 | 68 | 5.02 | 56 | 5.49 | 57 |
|  | 6 | 0.00 | 171 | 0.25 | 57 | 3.93 | 59 | 4.17 | 59 |
|  | 7 | 0.00 | 147 | 0.16 | 83 | 0.57 | 72 | 0.73 | 75 |
|  | 8+ | 0.00 | 210 |  |  | 1.00 | 64 | 1.00 | 64 |
|  | Total | 0.01 |  | 8.85 |  | 45.54 |  | 54.40 |  |
|  | SOP |  | 2 |  | 539 |  | 2,488 |  | 3,029 |
| Quarter |  | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 0.01 | 22 | 0.17 | 21 | 0.47 | 21 | 0.65 | 21 |
|  | 1 | 0.33 | 51 | 3.37 | 54 | 10.37 | 52 | 14.06 | 53 |
|  | 2 | 1.11 | 79 | 4.44 | 78 | 14.50 | 76 | 20.04 | 77 |
|  | 3 | 1.94 | 112 | 3.82 | 97 | 13.62 | 98 | 19.37 | 99 |
|  | 4 | 1.07 | 130 | 2.51 | 111 | 10.03 | 102 | 13.61 | 106 |
|  | 5 | 0.32 | 143 | 0.72 | 98 | 4.01 | 83 | 5.05 | 89 |
|  | 6 | 0.16 | 171 | 0.28 | 115 | 2.01 | 95 | 2.44 | 102 |
|  | 7 | 0.15 | 147 | 0.10 | 139 | 0.49 | 143 | 0.74 | 143 |
|  | 8+ | 0.02 | 210 | 0.07 | 171 | 0.51 | 111 | 0.61 | 122 |
|  | Total | 5.11 |  | 15.48 |  | 56.00 |  | 76.58 |  |
|  | SOP |  | 561 |  | 1,307 |  | 4,662 |  | 6,531 |
|  |  | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 0.01 | 22 | 0.17 | 21 | 0.47 | 21 | 0.65 | 21 |
|  | 1 | 11.07 | 15 | 6.24 | 48 | 27.45 | 38 | 44.76 | 34 |
|  | 2 | 13.69 | 49 | 7.17 | 70 | 51.22 | 57 | 72.07 | 57 |
|  | 3 | 13.87 | 87 | 8.01 | 87 | 97.12 | 83 | 119.00 | 84 |
|  | 4 | 9.07 | 120 | 5.81 | 96 | 86.85 | 101 | 101.73 | 102 |
|  | 5 | 4.22 | 147 | 1.84 | 104 | 36.94 | 124 | 43.00 | 125 |
|  | 6 | 2.60 | 161 | 0.83 | 113 | 27.94 | 143 | 31.36 | 144 |
|  | 7 | 1.63 | 166 | 0.61 | 142 | 19.87 | 178 | 22.11 | 176 |
|  | 8+ | 0.85 | 197 | 0.30 | 175 | 11.01 | 168 | 12.16 | 170 |
|  | Total | 57.00 |  | 30.97 |  | 358.87 |  | 446.84 |  |
|  | SOP |  | 4,612 |  | 2,479 |  | 34,770 |  | 41,861 |

Table 3.2.4 WESTERN BALTIC HERRING.
Samples of commercial landings by quarter and area for 2006 available to the Working Group.

|  | Country | Quarter | $\begin{array}{r} \text { Landings } \\ \text { in ' } 000 \text { tons } \\ \hline \end{array}$ | Numbers of samples | Numbers of fish meas. | $\begin{array}{r} \hline \begin{array}{l} \text { Numbers of } \\ \text { fish aged } \\ \hline \end{array} \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Skagerrak | Denmark | 1 | 882.0 | 14 | 1635 | 1581 |
|  |  | 2 | 14.0 | 1 | 1 | 0 |
|  |  | 3 | 3815.0 | 14 | 1320 | 1318 |
|  |  | 4 | 445.0 | 1 | 98 | 0 |
|  | Total |  | 5156.0 | 30 | 3,054 | 2,899 |
|  | Germany | 1 | 0.0 |  |  |  |
|  |  | 2 | 0.0 |  |  |  |
|  |  | 3 | 556.4 |  | No data available |  |
|  |  | 4 | 0.0 |  |  |  |
|  | Total |  | 556.4 | 0 | 0 | 0 |
|  | Faroe Islands | 1 |  |  |  |  |
|  |  | 2 |  |  |  |  |
|  |  | 3 |  |  |  |  |
|  |  | 4 |  |  |  |  |
|  | Total |  | 0.0 | 0 | 0 | 0 |
|  | Sweden | 1 | 8302.0 | 14 | 667 | 667 |
|  |  | 2 | 1782.0 | 8 | 700 | 700 |
|  |  | 3 | 9794.0 | 14 | 693 | 693 |
|  |  | 4 | 6167.0 | 12 | 669 | 669 |
|  | Total |  | 26045.0 | 48 | 2,729 | 2,729 |
| Kattegat | Denmark | 1 | 6058.0 | 27 | 3,061 | 3,010 |
|  |  | 2 | 330.0 | 4 | 178 | 24 |
|  |  | 3 | 1298.0 | 10 | 775 | 775 |
|  |  | 4 | 931.0 | 6 | 709 | 706 |
|  | Total |  | 8617.0 | 47.0 | 4723.0 | 4515.0 |
|  | Sweden | 1 | 5916.0 | 14 | 684 | 684 |
|  |  | 2 | 581.0 | 1 | 167 | 167 |
|  |  | 3 | 1061.0 | 6 | 662 | 662 |
|  |  | 4 | 3208.0 | 8 | 662 | 662 |
|  | Total |  | 10766.0 | 29 | 2,175 | 2,175 |
| Sub-Division 22 | Denmark | 1 | 1400.0 | 11 | 1,369 | 1,270 |
|  |  | 2 | 5.0 | No data available |  |  |
|  |  | 3 | 0.2 |  |  |  |
|  |  | 4 | 0.0 |  |  |  |
|  | Total |  | 1405.2 | 11 | 1,369 | 1,270 |
|  | Germany | 1 | 2036.5 | 3 | 1,330 | 358 |
|  |  | 2 | 608.2 | 1 | 323 | 60 |
|  |  | 3 | 1.6 | 0 | 0 | 0 |
|  |  | 4 | 561.2 | 0 | 0 | 0 |
|  | Total |  | 3207.5 | 4.0 | 1653.0 | 418.0 |
| Sub-Division 23 | Denmark | 1 | 619.0 | No data available |  |  |
|  |  | 2 | 13.0 |  |  |  |
|  |  | 3 | 306.0 |  |  |  |
|  |  | 4 | 889.0 |  |  |  |
|  | Total |  | 1827.0 |  |  |  |
|  | Sweden | 1 | 232.0 |  |  |  |
|  |  | 2 | 418.0 | No data available |  |  |
|  |  | 3 | 0.0 |  |  |  |
|  |  | 4 | 0.0 |  |  |  |
|  | Total |  | 650.0 | 0 | 0 | 0 |
| Sub-Division 24 | Denmark | 1 | 0.0 | No data available |  |  |
|  |  | 2 | 0.0 |  |  |  |
|  |  | 3 | 0.0 |  |  |  |
|  |  | 4 | 0.0 |  |  |  |
|  | Total |  | 0.0 | 0 | 0 | 0 |
|  | Germany | 1 | 6409.7 | 10 | 4,158 | 992 |
|  |  | 2 | 12420.5 | 22 | 9,805 | 1,486 |
|  |  | 3 | 0 | 0 | 0 | 0 |
|  |  | 4 | 832.0 | 1 | 426 | 113 |
|  | Total |  | 19662.3 | 33 | 14,389 | 2,591 |
|  | Poland | 1 | 936.3 | 1 | 472 | 89 |
|  |  | 2 | 2586.9 | 6 | 2,033 | 477 |
|  |  | 3 | 1545.4 | 2 | 643 | 215 |
|  |  | 4 | 419.2 | 1 | 316 | 102 |
|  | Total |  | 5487.8 | 10 | 3464 | 883 |
|  | Sweden | 1 | 3658.0 | 7 | 434 | 434 |
|  |  | 2 | 1606.0 | 5 | 288 | 288 |
|  |  | 3 | 934.0 | 4 | 395 | 395 |
|  |  | 4 | 3406.0 | 8 | 450 | 450 |
|  | Total |  | 9604.0 | 24 | 1,567 | 1,567 |

Table 3.2.5 WESTERN BALTIC HERRING.
Samples of landings by quarter and area used to estimate catch in numbers and mean weight by age for 2006.

|  | Country | Quarter | Fleet | Sampling |
| :---: | :---: | :---: | :---: | :---: |
| Skagerrak | Denmark | 1 | C | Danish sampling in Q1 |
|  |  | 2 | C | No landings |
|  |  | 3 | C | Danish sampling in Q3 |
|  |  | 4 | C | Danish sampling in Q4 |
|  | Germany | 1 | C | No landings |
|  |  | 2 | C | No landings |
|  |  | 3 | C | Danish sampling in Q3 |
|  |  | 4 | C | No landings |
|  | Sweden | 1 | C | Swedish sampling in Q1 |
|  |  | 2 | C | Swedish sampling in Q2 |
|  |  | 3 | C | Swedish sampling in Q3 |
|  |  | 4 | C | Swedish sampling in Q4 |
|  | Denmark | 1 | D | Danish sampling in Q1 |
|  |  | 2 | D | Danish sampling in Q1 |
|  |  | 3 | D | Danish sampling in Q3 |
|  |  | 4 | D | Danish sampling in Q3 |
|  | Sweden | 1 | D | Swedish sampling in Q1 |
|  |  | 2 | D | Swedish sampling in Q2 |
|  |  | 3 | D | Swedish sampling in Q3 |
|  |  | 4 | D | Swedish sampling in Q4 |
|  | Faroe Islands | 1 | C | No landings |
|  |  | 2 | C | No landings |
|  |  | 3 | C | No landings |
|  |  | 4 | C | No landings |
| Kattegat | Denmark | 1 | C | Danish sampling in Q1 |
|  |  | 2 | C | Danish sampling in Q2 |
|  |  | 3 | C | Danish sampling in Q3 |
|  |  | 4 | C | Danish sampling in Q4 |
|  | Sweden | 1 | C | Swedish sampling in Q1 |
|  |  | 2 | C | Swedish sampling in Q2 |
|  |  | 3 | C | Swedish sampling in Q3 |
|  |  | 4 | C | Swedish sampling in Q4 |
|  | Denmark | 1 | D | Danish sampling in Q1 |
|  |  | 2 | D | Danish sampling in Q1 |
|  |  | 3 | D | Danish sampling in Q3 |
|  |  | 4 | D | Danish sampling in Q4 |
|  | Sweden | 1 | D | Swedish sampling in Q1 |
|  |  | 2 | D | No landings |
|  |  | 3 | D | Swedish sampling in Q3 |
|  |  | 4 | D | Swedish sampling in Q4 |

Fleet $\mathbf{C}=$ Human consumption, Fleet $\mathbf{D}=$ Industrial landings.
continued
Table 3.2.5 WESTERN BALTIC HERRING.
Samples of landings by quarter and area used to estimate catch in numbers and mean weight by age for 2006

|  | Country | Quarter | Fleet | Sampling |
| :---: | :---: | :---: | :---: | :---: |
| Sub-Division 22 | Denmark | 1 | F | Danish sampling in Q1 |
|  |  | 2 | F | Danish sampling in Q1 |
|  |  | 3 | F | No landings |
|  |  | 4 | F | Swedish sampling in Q4 in Sub-division 24 |
|  | Germany | 1 | F | German sampling in Q1 |
|  |  | 2 | F | German sampling in Q2 |
|  |  | 3 | F | German sampling in Q4 |
|  |  | 4 | F | German sampling in Q4 |
| Sub-Division 23 | Denmark | 1 | F | Swedish sampling in Q1 in Sub-division 24 |
|  |  | 2 | F | Swedish sampling in Q2 in Sub-division 24 |
|  |  | 3 | F | Swedish sampling in Q3 in Sub-division 24 |
|  |  | 4 | F | Swedish sampling in Q4 in Sub-division 24 |
|  | Sweden | 1 | F | Danish sampling in Q1 in Sub-division 22 |
|  |  | 2 | F | Swedish sampling in Q2 in Sub-division 24 |
|  |  | 3 | F | Swedish sampling in Q3 in Sub-division 24 |
|  |  | 4 | F | Swedish sampling in Q4 in Sub-division 24 |
| Sub-Division 24 | Denmark | 1 | F | No landings |
|  |  | 2 | F | No landings |
|  |  | 3 | F | No landings |
|  |  | 4 | F | No landings |
|  | Germany | 1 | F | German sampling in Q1 |
|  |  | 2 | F | German sampling in Q2 |
|  |  | 3 | F | Swedish sampling in Q3 |
|  |  | 4 | F | German sampling in Q4 |
|  | Poland | 1 | F | Polish sampling in Q1 |
|  |  | 2 | F | Polish sampling in Q2 |
|  |  | 3 | F | Polish sampling in Q3 |
|  |  | 4 | F | Polish sampling in Q4 |
|  | Sweden | 1 | F | Swedish sampling in Q1 |
|  |  | 2 | F | Swedish sampling in Q2 |
|  |  | 3 | F | Swedish sampling in Q3 |
|  |  | 4 | F | Swedish sampling in Q4 |

Fleet $\mathrm{C}=$ Human consumption, Fleet $\mathrm{D}=$ Industrial landings, Fleet $\mathrm{E}=$ All landings from sub.div.22-24.

Table 3.2.6 WESTERN BALTIC HERRING.
Proportion of North Sea autumn spawners and Western Baltic spring spawners given in \% in Skagerrak and Kattegat by age and quarter.


Age-classes with few otolith analyses were supplemented with analyses from acoustic survey sampling and/or pooled into plus-groups with more than 11 individuals as indicated by bold figures and in bracketts in the source column.
$a=$ supplemented with acoustic samples, $s k=$ assumed equal to Skagerrak

Table 3.2.7 WESTERN BALTIC HERRING. North Sea autumn spawners in Kattegat. Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.

North Sea Autumn spawners

|  | Division: |  | Kattegat | Year: | 2006 | Country: All |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 8.05 | 31 | 35.42 | 17 | 43.47 | 20 |
|  | 2 | 11.35 | 68 | 2.11 | 49 | 13.45 | 65 |
|  | 3 | 0.66 | 97 | 0.04 | 102 | 0.70 | 97 |
|  | 4 | 0.17 | 118 | 0.01 | 122 | 0.18 | 118 |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | 8+ |  |  |  |  |  |  |
|  | Total | 20.23 |  | 37.58 |  | 57.81 |  |
|  | SOP |  | 1,109 |  | 711 |  | 1,821 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 6.82 | 48 | 0.42 | 17 | 7.24 | 47 |
|  | 2 | 2.10 | 77 | 0.01 | 35 | 2.11 | 76 |
|  | 3 | 0.02 | 102 | 0.00 | 144 | 0.02 | 102 |
|  | 4 |  |  |  |  |  |  |
|  | 5 | 0.00 | 144 |  |  | 0.00 | 144 |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | $8+$ | 0.00 | 171 |  |  | 0.00 | 171 |
|  | Total | 8.95 |  | 0.43 |  | 9.38 |  |
|  | SOP |  | 494 |  | 7 |  | 502 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 |  |  | 21.86 | 12 | 21.86 | 12 |
|  | 1 | 2.37 | 66 | 1.75 | 42 | 4.12 | 56 |
|  | 2 | 0.11 | 78 | 0.00 | 77 | 0.12 | 78 |
|  | 3 | 0.08 | 93 | 0.00 | 99 | 0.08 | 94 |
|  | 4 | 0.05 | 107 | 0.00 | 112 | 0.05 | 107 |
|  | 5 |  |  |  |  |  |  |
|  | 6 | 0.01 | 163 | 0.00 | 163 | 0.01 | 163 |
|  | 7 | 0.00 | 153 | 0.00 | 153 | 0.00 | 153 |
|  | 8+ |  |  |  |  |  |  |
|  | Total | 2.62 |  | 23.61 |  | 26.23 |  |
|  | SOP |  | 180 |  | 341 |  | 521 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 1.77 | 29 | 1.19 | 15 | 2.96 | 23 |
|  | 1 | 11.58 | 58 | 2.46 | 61 | 14.04 | 59 |
|  | 2 | 0.61 | 86 | 0.14 | 89 | 0.75 | 86 |
|  | 3 | 0.03 | 122 | 0.01 | 127 | 0.03 | 123 |
|  | 4 | 0.03 | 145 | 0.01 | 147 | 0.03 | 145 |
|  | 5 |  |  |  |  |  |  |
|  | 6 | 0.01 | 191 | 0.00 | 191 | 0.01 | 191 |
|  | 7 | 0.01 | 231 | 0.00 | 231 | 0.02 | 231 |
|  | 8+ | 0.00 | 215 | 0.00 | 215 | 0.01 | 215 |
|  | Total | 14.04 |  | 3.81 |  | 17.85 |  |
|  | SOP |  | 791 |  | 182 |  | 973 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 1.77 | 29 | 23.05 | 12 | 24.82 | 13 |
|  | 1 | 28.83 | 49 | 40.05 | 21 | 68.87 | 33 |
|  | 2 | 14.18 | 70 | 2.26 | 52 | 16.43 | 68 |
|  | 3 | 0.79 | 97 | 0.05 | 105 | 0.84 | 98 |
|  | 4 | 0.25 | 118 | 0.01 | 133 | 0.26 | 119 |
|  | 5 | 0.00 | 144 |  |  | 0.00 | 144 |
|  | 6 | 0.02 | 177 | 0.00 | 189 | 0.02 | 179 |
|  | 7 | 0.01 | 225 | 0.00 | 231 | 0.02 | 226 |
|  | 8+ | 0.01 | 202 | 0.00 | 215 | 0.01 | 204 |
|  | Total | 45.84 |  | 65.43 |  | 111.27 |  |
|  | SOP |  | 2,575 |  | 1,242 |  | 3,817 |

Table 3.2.8 WESTERN BALTIC HERRING. North Sea autumn spawners in Skagerrak. Landings in numbers (mill.), mean weight (g.) and SOP (t) by age,
quarter and fleet.
North Sea Autumn spawners

|  | Division: |  | Skagerrak | Year: | 2006 | Country: | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 0.87 | 31 | 5.04 | 17 | 5.92 | 19 |
|  | 2 | 8.24 | 77 | 4.82 | 74 | 13.06 | 76 |
|  | 3 | 2.97 | 109 | 2.43 | 109 | 5.40 | 109 |
|  | 4 | 0.75 | 131 | 0.62 | 131 | 1.37 | 131 |
|  | 5 | 1.24 | 187 | 1.05 | 187 | 2.29 | 187 |
|  | 6 | 0.15 | 195 | 0.13 | 195 | 0.27 | 195 |
|  | 7 | 0.05 | 220 | 0.05 | 220 | 0.10 | 220 |
|  | 8+ | 0.04 | 209 | 0.03 | 209 | 0.07 | 209 |
|  | Total | 14.32 |  | 14.17 |  | 28.49 |  |
|  | SOP |  | 1,365 |  | 1,024 |  | 2,389 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 2.58 | 41 | 0.18 | 41 | 2.76 | 41 |
|  | 2 | 8.64 | 75 | 0.59 | 75 | 9.22 | 75 |
|  | 3 | 0.01 | 105 | 0.00 | 105 | 0.01 | 105 |
|  | 4 | 0.02 | 135 | 0.00 | 135 | 0.02 | 135 |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | 8+ |  |  |  |  |  |  |
|  | Total | 11.25 |  | 0.77 |  | 12.01 |  |
|  | SOP |  | 757 |  | 52 |  | 809 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.39 | 38 | 4.73 | 16 | 5.12 | 18 |
|  | 1 | 34.04 | 83 | 2.67 | 47 | 36.71 | 81 |
|  | 2 | 9.94 | 103 | 0.11 | 99 | 10.05 | 103 |
|  | 3 | 2.44 | 131 | 0.02 | 129 | 2.46 | 131 |
|  | 4 | 0.79 | 148 | 0.01 | 147 | 0.80 | 148 |
|  | 5 | 0.32 | 162 | 0.01 | 177 | 0.33 | 163 |
|  | 6 | 0.20 | 175 | 0.00 | 167 | 0.20 | 175 |
|  | 7 | 0.11 | 188 | 0.00 | 187 | 0.11 | 188 |
|  | $8+$ | 0.04 | 204 | 0.00 | 219 | 0.04 | 204 |
|  | Total | 48.27 |  | 7.55 |  | 55.81 |  |
|  | SOP |  | 4,422 |  | 218 |  | 4,641 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 3.88 | 15 | 1.27 | 15 | 5.15 | 15 |
|  | 1 | 26.99 | 73 | 8.88 | 73 | 35.87 | 73 |
|  | 2 | 1.06 | 104 | 0.35 | 104 | 1.41 | 104 |
|  | 3 | 1.12 | 140 | 0.37 | 140 | 1.48 | 140 |
|  | 4 | 0.61 | 155 | 0.20 | 155 | 0.82 | 155 |
|  | 5 | 0.54 | 192 | 0.18 | 192 | 0.72 | 192 |
|  | 6 | 0.06 | 216 | 0.02 | 216 | 0.08 | 216 |
|  | 7 | 0.11 | 232 | 0.04 | 232 | 0.15 | 232 |
|  | $8+$ | 0.04 | 207 | 0.01 | 207 | 0.06 | 207 |
|  | Total | 34.41 |  | 11.32 |  | 45.73 |  |
|  | SOP |  | - 2,528 |  | 831 |  | 3,360 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 4.26 | 17 | 6.01 | 16 | 10.27 | 16 |
|  | 1 | 64.49 | 76 | 16.77 | 51 | 81.25 | 71 |
|  | 2 | 27.88 | 87 | 5.87 | 76 | 33.74 | 85 |
|  | 3 | 6.54 | 122 | 2.82 | 113 | 9.36 | 119 |
|  | 4 | 2.17 | 144 | 0.82 | 137 | 3.00 | 142 |
|  | 5 | 2.11 | 184 | 1.24 | 188 | 3.34 | 186 |
|  | 6 | 0.40 | 188 | 0.15 | 197 | 0.55 | 191 |
|  | 7 | 0.28 | 212 | 0.08 | 225 | 0.36 | 215 |
|  | $8+$ | 0.12 | 206 | 0.05 | 208 | 0.17 | 207 |
|  | Total | 108.25 |  | 33.79 |  | 142.04 |  |
|  | SOP |  | 9,073 |  | 2,125 |  | 11,198 |

Table 3.2.9 WESTERN BALTIC HERRING. Spring Spawners in Kattegat. Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.

Western Baltic Spring spawners

|  | Division: |  | Kattegat | Year: | 2006 | Country: All |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 7.22 | 31 | 31.79 | 17 | 39.02 | 20 |
|  | 2 | 23.47 | 68 | 4.36 | 49 | 27.82 | 65 |
|  | 3 | 37.07 | 97 | 2.50 | 102 | 39.56 | 97 |
|  | 4 | 8.99 | 118 | 0.38 | 122 | 9.37 | 118 |
|  | 5 | 11.81 | 174 | 0.63 | 170 | 12.45 | 174 |
|  | 6 | 1.26 | 183 | 0.06 | 189 | 1.32 | 183 |
|  | 7 | 0.90 | 171 | 0.02 | 147 | 0.92 | 171 |
|  | 8+ | 0.37 | 171 | 0.02 | 167 | 0.39 | 171 |
|  | Total | 91.09 |  | 39.76 |  | 130.85 |  |
|  | SOP |  | 8,972 |  | 1,181 |  | 10,152 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 |  |  |  |  |  |  |
|  | 2 | 3.12 | 77 | 0.01 | 35 | 3.14 | 76 |
|  | 3 | 1.04 | 102 | 0.00 | 144 | 1.04 | 102 |
|  | 4 | 0.37 | 99 |  |  | 0.37 | 99 |
|  | 5 | 0.12 | 144 |  |  | 0.12 | 144 |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | 8+ | 0.06 | 171 |  |  | 0.06 | 171 |
|  | Total | 4.71 |  | 0.02 |  | 4.72 |  |
|  | SOP |  | 409 |  | 1 |  | 409 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 |  |  | 0.26 | 12 | 0.26 | 12 |
|  | 1 | 5.32 | 66 | 3.92 | 42 | 9.24 | 56 |
|  | 2 | 7.48 | 78 | 0.15 | 77 | 7.63 | 78 |
|  | 3 | 4.80 | 93 | 0.06 | 99 | 4.87 | 94 |
|  | 4 | 1.76 | 107 | 0.02 | 112 | 1.78 | 107 |
|  | 5 | 0.42 | 120 | 0.01 | 133 | 0.43 | 120 |
|  | 6 | 0.12 | 163 | 0.00 | 163 | 0.12 | 163 |
|  | 7 | 0.02 | 153 | 0.00 | 153 | 0.02 | 153 |
|  | 8+ |  |  |  |  |  |  |
|  | Total | 19.93 |  | 4.42 |  | 24.35 |  |
|  | SOP |  | 1,648 |  | 190 |  | 1,838 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 1.74 | 29 | 1.17 | 15 | 2.90 | 23 |
|  | 1 | 23.53 | 58 | 4.99 | 61 | 28.52 | 59 |
|  | 2 | 6.72 | 86 | 1.52 | 89 | 8.24 | 86 |
|  | 3 | 2.69 | 122 | 0.67 | 127 | 3.36 | 123 |
|  | 4 | 0.88 | 145 | 0.24 | 147 | 1.12 | 145 |
|  | 5 | 0.26 | 172 | 0.07 | 172 | 0.33 | 172 |
|  | 6 | 0.09 | 191 | 0.03 | 191 | 0.12 | 191 |
|  | 7 | 0.14 | 231 | 0.04 | 231 | 0.18 | 231 |
|  | 8+ | 0.05 | 215 | 0.01 | 215 | 0.06 | 215 |
|  | Total | 36.10 |  | 8.74 |  | 44.84 |  |
|  | SOP |  | 2,561 |  | 606 |  | 3,166 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 1.74 | 29 | 1.43 | 14 | 3.17 | 22 |
|  | 1 | 36.07 | 54 | 40.70 | 25 | 76.78 | 39 |
|  | 2 | 40.79 | 73 | 6.05 | 60 | 46.84 | 72 |
|  | 3 | 45.60 | 98 | 3.23 | 107 | 48.83 | 99 |
|  | 4 | 12.00 | 117 | 0.64 | 131 | 12.64 | 118 |
|  | 5 | 12.61 | 172 | 0.71 | 170 | 13.32 | 172 |
|  | 6 | 1.47 | 182 | 0.09 | 188 | 1.57 | 182 |
|  | 7 | 1.06 | 179 | 0.06 | 202 | 1.12 | 180 |
|  | 8+ | 0.47 | 175 | 0.03 | 186 | 0.51 | 176 |
|  | Total | 151.82 |  | 52.94 |  | 204.75 |  |
|  | SOP |  | 13,589 |  | 1,977 |  | 15,566 |

Table 3.2.10 WESTERN BALTIC HERRING. Spring spawners in Skagerrak. Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.

Western Baltic Spring spawners

|  | Division: |  | Skagerrak | Year: | 2006 | Country: All |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 1.08 | 31 | 6.25 | 17 | 7.33 | 19 |
|  | 2 | 16.13 | 77 | 9.43 | 74 | 25.56 | 76 |
|  | 3 | 10.90 | 109 | 8.91 | 109 | 19.81 | 109 |
|  | 4 | 2.75 | 131 | 2.26 | 131 | 5.01 | 131 |
|  | 5 | 4.55 | 187 | 3.85 | 187 | 8.40 | 187 |
|  | 6 | 0.54 | 195 | 0.46 | 195 | 1.00 | 195 |
|  | 7 | 0.20 | 220 | 0.17 | 220 | 0.36 | 220 |
|  | 8+ | 0.15 | 209 | 0.13 | 209 | 0.27 | 209 |
|  | Total | 36.29 |  | 31.45 |  | 67.74 |  |
|  | SOP |  | 3,852 |  | 2,940 |  | 6,793 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 0.22 | 41 | 0.02 | 41 | 0.24 | 41 |
|  | 2 | 3.24 | 75 | 0.22 | 75 | 3.46 | 75 |
|  | 3 | 2.93 | 105 | 0.20 | 105 | 3.13 | 105 |
|  | 4 | 0.75 | 135 | 0.05 | 135 | 0.80 | 135 |
|  | 5 | 1.22 | 170 | 0.24 | 118 | 1.46 | 162 |
|  | 6 | 0.03 | 179 | 0.00 | 179 | 0.03 | 179 |
|  | 7 | 0.11 | 179 | 0.01 | 179 | 0.12 | 179 |
|  | 8+ | 0.09 | 184 | 0.01 | 184 | 0.09 | 184 |
|  | Total | 8.59 |  | 0.74 |  | 9.33 |  |
|  | SOP |  | 911 |  | 76 |  | 988 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 |  |  |  |  |  |  |
|  | 1 | 2.52 | 83 | 0.20 | 47 | 2.72 | 81 |
|  | 2 | 26.09 | 103 | 0.29 | 99 | 26.38 | 103 |
|  | 3 | 22.10 | 131 | 0.16 | 129 | 22.26 | 131 |
|  | 4 | 11.06 | 148 | 0.07 | 147 | 11.14 | 148 |
|  | 5 | 5.49 | 162 | 0.12 | 177 | 5.60 | 163 |
|  | 6 | 2.54 | 175 | 0.02 | 167 | 2.56 | 175 |
|  | 7 | 2.51 | 188 | 0.02 | 187 | 2.53 | 188 |
|  | 8+ | 0.93 | 204 | 0.01 | 219 | 0.94 | 204 |
|  | Total | 73.25 |  | 0.88 |  | 74.14 |  |
|  | SOP |  | 9,419 |  | 98 |  | 9,518 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 0.37 | 15 | 0.12 | 15 | 0.49 | 15 |
|  | 1 | 1.86 | 73 | 0.61 | 73 | 2.47 | 73 |
|  | 2 | 4.76 | 104 | 1.57 | 104 | 6.33 | 104 |
|  | 3 | 5.02 | 140 | 1.65 | 140 | 6.67 | 140 |
|  | 4 | 2.77 | 155 | 0.91 | 155 | 3.68 | 155 |
|  | 5 | 2.44 | 192 | 0.80 | 192 | 3.25 | 192 |
|  | 6 | 0.26 | 216 | 0.08 | 216 | 0.34 | 216 |
|  | 7 | 0.51 | 232 | 0.17 | 232 | 0.68 | 232 |
|  | 8+ | 0.19 | 207 | 0.06 | 207 | 0.26 | 207 |
|  | Total | 18.18 |  | 5.98 |  | 24.16 |  |
|  | SOP |  | 2,448 |  | 805 |  | 3,252 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 0.37 | 15 | 0.12 | 15 | 0.49 | 15 |
|  | 1 | 5.69 | 68 | 7.07 | 23 | 12.76 | 43 |
|  | 2 | 50.22 | 93 | 11.50 | 79 | 61.72 | 90 |
|  | 3 | 40.95 | 124 | 10.93 | 114 | 51.88 | 122 |
|  | 4 | 17.33 | 146 | 3.29 | 138 | 20.62 | 145 |
|  | 5 | 13.70 | 176 | 5.01 | 184 | 18.71 | 178 |
|  | 6 | 3.37 | 182 | 0.56 | 197 | 3.94 | 184 |
|  | 7 | 3.34 | 196 | 0.37 | 222 | 3.70 | 199 |
|  | 8+ | 1.36 | 203 | 0.20 | 208 | 1.56 | 204 |
|  | Total | 136.32 |  | 39.05 |  | 175.38 |  |
|  | SOP |  | 16,631 |  | 3,920 |  | 20,551 |

Table 3.2.11 WESTERN BALTIC HERRING. Autumn Spawners in Division Illa. Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet. North Sea Autumn spawners

| Quarter | Division: |  | IIIa | Year: | 2006 | Country: All |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 8.92 | 31 | 40.47 | 17 | 49.39 | 20 |
|  | 2 | 19.59 | 72 | 6.93 | 66 | 26.52 | 70 |
|  | 3 | 3.63 | 106 | 2.48 | 109 | 6.11 | 107 |
|  | 4 | 0.92 | 128 | 0.62 | 131 | 1.55 | 129 |
|  | 5 | 1.24 | 187 | 1.05 | 187 | 2.29 | 187 |
|  | 6 | 0.15 | 195 | 0.13 | 195 | 0.27 | 195 |
|  | 7 | 0.05 | 220 | 0.05 | 220 | 0.10 | 220 |
|  | 8+ | 0.04 | 209 | 0.03 | 209 | 0.07 | 209 |
|  | Total | 34.55 |  | 51.75 |  | 86.30 |  |
|  | SOP |  | 2,475 |  | 1,735 |  | 4,210 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 9.40 | 46 | 0.59 | 24 | 9.99 | 45 |
|  | 2 | 10.74 | 75 | 0.60 | 74 | 11.34 | 75 |
|  | 3 | 0.03 | 103 | 0.00 | 106 | 0.03 | 103 |
|  | 4 | 0.02 | 135 | 0.00 | 135 | 0.02 | 135 |
|  | 5 | 0.00 | 144 |  |  | 0.00 | 144 |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | 8+ | 0.00 | 171 |  |  | 0.00 | 171 |
|  | Total | 20.20 |  | 1.19 |  | 21.39 |  |
|  | SOP |  | 1,252 |  | 59 |  | 1,311 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.39 | 38 | 26.59 | 13 | 26.98 | 13 |
|  | 1 | 36.41 | 82 | 4.42 | 45 | 40.83 | 78 |
|  | 2 | 10.05 | 103 | 0.11 | 99 | 10.16 | 103 |
|  | 3 | 2.52 | 129 | 0.02 | 128 | 2.54 | 129 |
|  | 4 | 0.84 | 146 | 0.01 | 143 | 0.84 | 146 |
|  | 5 | 0.32 | 162 | 0.01 | 177 | 0.33 | 163 |
|  | 6 | 0.20 | 175 | 0.00 | 167 | 0.20 | 175 |
|  | 7 | 0.11 | 188 | 0.00 | 187 | 0.11 | 188 |
|  | 8+ | 0.04 | 204 | 0.00 | 219 | 0.04 | 204 |
|  | Total | 50.89 |  | 31.16 |  | 82.05 |  |
|  | SOP |  | 4,602 |  | 559 |  | 5,162 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 5.64 | 19 | 2.47 | 15 | 8.11 | 18 |
|  | 1 | 38.58 | 68 | 11.33 | 70 | 49.91 | 69 |
|  | 2 | 1.67 | 97 | 0.49 | 100 | 2.16 | 98 |
|  | 3 | 1.14 | 139 | 0.37 | 139 | 1.52 | 139 |
|  | 4 | 0.64 | 155 | 0.21 | 155 | 0.85 | 155 |
|  | 5 | 0.54 | 192 | 0.18 | 192 | 0.72 | 192 |
|  | 6 | 0.07 | 213 | 0.02 | 213 | 0.09 | 213 |
|  | 7 | 0.13 | 232 | 0.04 | 232 | 0.17 | 232 |
|  | 8+ | 0.05 | 207 | 0.02 | 207 | 0.06 | 207 |
|  | Total | 48.45 |  | 15.12 |  | 63.58 |  |
|  | SOP |  | 3,319 |  | 1,014 |  | 4,333 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 6.03 | 20 | 29.06 | 13 | 35.09 | 14 |
|  | 1 | 93.31 | 68 | 56.82 | 30 | 150.13 | 54 |
|  | 2 | 42.05 | 81 | 8.12 | 69 | 50.18 | 79 |
|  | 3 | 7.33 | 119 | 2.87 | 113 | 10.20 | 118 |
|  | 4 | 2.42 | 141 | 0.84 | 137 | 3.26 | 140 |
|  | 5 | 2.11 | 184 | 1.24 | 188 | 3.34 | 186 |
|  | 6 | 0.42 | 188 | 0.15 | 197 | 0.56 | 190 |
|  | 7 | 0.29 | 213 | 0.09 | 225 | 0.38 | 216 |
|  | 8+ | 0.13 | 206 | 0.05 | 209 | 0.18 | 207 |
|  | Total | 154.09 |  | 99.22 |  | 253.31 |  |
|  | SOP |  | 11,648 |  | 3,367 |  | 15,015 |

Table 3.2.12 WESTERN BALTIC HERRING. Spring spawners in Division Illa. Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.

Western Baltic Spring spawners

| Quarter | Division: |  | Illa | Year: | 2006 | Country: All |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 8.30 | 31 | 38.04 | 17 | 46.34 | 20 |
|  | 2 | 39.59 | 72 | 13.79 | 66 | 53.38 | 70 |
|  | 3 | 47.97 | 99 | 11.41 | 107 | 59.38 | 101 |
|  | 4 | 11.74 | 121 | 2.64 | 130 | 14.38 | 122 |
|  | 5 | 16.36 | 178 | 4.48 | 185 | 20.84 | 179 |
|  | 6 | 1.80 | 186 | 0.52 | 194 | 2.32 | 188 |
|  | 7 | 1.09 | 180 | 0.19 | 212 | 1.28 | 185 |
|  | 8+ | 0.52 | 182 | 0.15 | 203 | 0.66 | 187 |
|  | Total | 127.38 |  | 71.21 |  | 198.59 |  |
|  | SOP |  | 12,824 |  | 4,121 |  | 16,945 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 0.22 | 41 | 0.02 | 41 | 0.24 | 41 |
|  | 2 | 6.36 | 76 | 0.23 | 73 | 6.60 | 76 |
|  | 3 | 3.98 | 105 | 0.20 | 106 | 4.18 | 105 |
|  | 4 | 1.11 | 123 | 0.05 | 135 | 1.17 | 124 |
|  | 5 | 1.33 | 168 | 0.24 | 118 | 1.57 | 161 |
|  | 6 | 0.03 | 179 | 0.00 | 179 | 0.03 | 179 |
|  | 7 | 0.11 | 179 | 0.01 | 179 | 0.12 | 179 |
|  | 8+ | 0.14 | 179 | 0.01 | 184 | 0.15 | 179 |
|  | Total | 13.30 |  | 0.76 |  | 14.05 |  |
|  | SOP |  | 1,320 |  | 77 |  | 1,397 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 |  |  | 0.26 | 12 | 0.26 | 12 |
|  | 1 | 7.84 | 72 | 4.12 | 43 | 11.96 | 62 |
|  | 2 | 33.57 | 97 | 0.44 | 91 | 34.01 | 97 |
|  | 3 | 26.90 | 124 | 0.22 | 121 | 27.13 | 124 |
|  | 4 | 12.83 | 142 | 0.09 | 139 | 12.92 | 142 |
|  | 5 | 5.91 | 159 | 0.12 | 175 | 6.03 | 160 |
|  | 6 | 2.67 | 175 | 0.02 | 166 | 2.68 | 175 |
|  | 7 | 2.53 | 188 | 0.02 | 187 | 2.55 | 188 |
|  | 8+ | 0.93 | 204 | 0.01 | 219 | 0.94 | 204 |
|  | Total | 93.18 |  | 5.31 |  | 98.48 |  |
|  | SOP |  | 11,068 |  | 289 |  | 11,356 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 2.10 | 26 | 1.29 | 15 | 3.40 | 22 |
|  | 1 | 25.39 | 59 | 5.60 | 62 | 30.99 | 60 |
|  | 2 | 11.48 | 93 | 3.09 | 97 | 14.57 | 94 |
|  | 3 | 7.71 | 134 | 2.32 | 136 | 10.02 | 134 |
|  | 4 | 3.65 | 153 | 1.15 | 153 | 4.80 | 153 |
|  | 5 | 2.70 | 190 | 0.88 | 190 | 3.58 | 190 |
|  | 6 | 0.35 | 209 | 0.11 | 210 | 0.46 | 209 |
|  | 7 | 0.66 | 232 | 0.21 | 232 | 0.86 | 232 |
|  | 8+ | 0.24 | 208 | 0.08 | 208 | 0.32 | 208 |
|  | Total | 54.28 |  | 14.72 |  | 69.00 |  |
|  | SOP |  | 5,008 |  | 1,410 |  | 6,419 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 2.10 | 26 | 1.55 | 14 | 3.66 | 21 |
|  | 1 | 41.76 | 56 | 47.78 | 24 | 89.54 | 39 |
|  | 2 | 91.01 | 84 | 17.55 | 72 | 108.56 | 82 |
|  | 3 | 86.55 | 110 | 14.15 | 112 | 100.71 | 111 |
|  | 4 | 29.33 | 134 | 3.93 | 137 | 33.26 | 135 |
|  | 5 | 26.31 | 174 | 5.72 | 182 | 32.03 | 176 |
|  | 6 | 4.85 | 182 | 0.65 | 196 | 5.50 | 183 |
|  | 7 | 4.39 | 192 | 0.43 | 219 | 4.82 | 195 |
|  | 8+ | 1.83 | 196 | 0.23 | 205 | 2.07 | 197 |
|  | Total | 288.14 |  | 91.99 |  | 380.13 |  |
|  | SOP |  | 30,220 |  | 5,897 |  | 36,116 |

Table 3.2.13 WESTERN BALTIC HERRING.
Total catch in numbers (mill) and mean weight (g), SOP (tonnes) of Western Baltic Spring spawners in Division IIIa and the North Sea in the years 1991-2006.

|  | W-rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year |  |  |  |  |  |  |  |  |  |  |  |
| 1991 | Numbers | 100.00 | 157.43 | 382.91 | 394.77 | 166.97 | 112.35 | 21.86 | 7.33 | 3.15 | 1,346.77 |
|  | Mean W. | 33.0 | 48.6 | 69.5 | 99.9 | 135.7 | 146.2 | 166.9 | 179.7 | 193.2 |  |
|  | SOP | 3,300 | 7,656 | 26,614 | 39,455 | 22,657 | 16,430 | 3,648 | 1,318 | 609 | 121,687 |
| 1992 | Numbers | 109.08 | 246.00 | 321.85 | 174.02 | 154.47 | 78.33 | 55.83 | 17.91 | 8.53 | 1,166.03 |
|  | Mean W. | 13.9 | 44.1 | 87.0 | 112.9 | 136.2 | 166.3 | 183.5 | 194.4 | 203.6 |  |
|  | SOP | 1,516 | 10,841 | 27,986 | 19,653 | 21,035 | 13,030 | 10,243 | 3,481 | 1,737 | 109,523 |
| 1993 | Numbers | 161.25 | 371.50 | 315.82 | 219.05 | 94.08 | 59.43 | 40.97 | 21.71 | 8.22 | 1,292.03 |
|  | Mean W. | 15.1 | 25.9 | 81.4 | 127.5 | 150.1 | 171.1 | 195.9 | 209.1 | 239.0 |  |
|  | SOP | 2,435 | 9,612 | 25,696 | 27,936 | 14,120 | 10,167 | 8,027 | 4,541 | 1,966 | 104,498 |
| 1994 | Numbers | 60.62 | 153.11 | 261.14 | 221.64 | 130.97 | 77.30 | 44.40 | 14.39 | 8.62 | 972.19 |
|  | Mean W. | 20.2 | 42.6 | 94.8 | 122.7 | 150.3 | 168.7 | 194.7 | 209.9 | 220.2 |  |
|  | SOP | 1,225 | 6,524 | 24,767 | 27,206 | 19,686 | 13,043 | 8,642 | 3,022 | 1,898 | 106,013 |
| 1995 | Numbers | 50.31 | 302.51 | 204.19 | 97.93 | 90.86 | 30.55 | 21.28 | 12.01 | 7.24 | 816.86 |
|  | Mean W. | 17.9 | 41.5 | 97.8 | 138.0 | 163.1 | 198.5 | 207.0 | 228.8 | 234.3 |  |
|  | SOP | 902 | 12,551 | 19,970 | 13,517 | 14,823 | 6,065 | 4,404 | 2,747 | 1,696 | 76,674 |
| 1996 | Numbers | 166.23 | 228.05 | 317.74 | 75.60 | 40.41 | 30.63 | 12.58 | 6.73 | 5.63 | 883.60 |
|  | Mean W. | 10.5 | 27.6 | 90.1 | 134.9 | 164.9 | 186.6 | 204.1 | 208.5 | 220.2 |  |
|  | SOP | 1,748 | 6,296 | 28,618 | 10,197 | 6,665 | 5,714 | 2,568 | 1,402 | 1,241 | 64,449 |
| 1997 | Numbers | 25.97 | 73.43 | 158.71 | 180.06 | 30.15 | 14.15 | 4.77 | 1.75 | 2.31 | 491.31 |
|  | Mean W. | 19.2 | 49.7 | 76.7 | 127.2 | 154.4 | 175.8 | 184.4 | 192.0 | 208.0 |  |
|  | SOP | 498 | 3,648 | 12,176 | 22,913 | 4,656 | 2,489 | 879 | 337 | 480 | 48,075 |
| 1998 | Numbers | 36.26 | 175.14 | 315.15 | 94.53 | 54.72 | 11.19 | 8.72 | 2.19 | 2.09 | 699.98 |
|  | Mean W. | 27.8 | 51.3 | 71.5 | 108.8 | 142.6 | 171.7 | 194.4 | 184.2 | 230.0 |  |
|  | SOP | 1,009 | 8,980 | 22,542 | 10,287 | 7,804 | 1,922 | 1,695 | 403 | 481 | 55,121 |
| 1999 | Numbers | 41.34 | 190.29 | 155.67 | 122.26 | 43.16 | 22.21 | 4.42 | 3.02 | 2.40 | 584.77 |
|  | Mean W. | 11.5 | 51.0 | 83.6 | 114.9 | 121.2 | 145.2 | 169.6 | 123.8 | 152.3 |  |
|  | SOP | 477 | 9,698 | 13,012 | 14,048 | 5,232 | 3,225 | 749 | 373 | 366 | 47,179 |
| 2000 | Numbers | 114.83 | 318.22 | 302.10 | 99.88 | 50.85 | 18.76 | 8.21 | 1.35 | 1.40 | 915.60 |
|  | Mean W. | 22.6 | 31.9 | 67.4 | 107.7 | 140.2 | 170.0 | 157.0 | 185.0 | 210.1 |  |
|  | SOP | 2,601 | 10,145 | 20,357 | 10,756 | 7,131 | 3,189 | 1,288 | 249 | 294 | 56,010 |
| 2001 | Numbers | 121.68 | 36.63 | 208.10 | 111.08 | 32.06 | 19.67 | 9.84 | 4.17 | 2.42 | 545.65 |
|  | Mean W. | 9.0 | 51.2 | 76.2 | 108.9 | 145.3 | 171.4 | 188.2 | 187.2 | 203.3 |  |
|  | SOP | 1,096 | 1,875 | 15,863 | 12,093 | 4,657 | 3,371 | 1,852 | 780 | 492 | 42,079 |
| 2002 | Numbers | 69.63 | 577.69 | 168.26 | 134.60 | 53.09 | 12.05 | 7.48 | 2.43 | 2.02 | 1,027.26 |
|  | Mean W. | 10.2 | 20.4 | 78.2 | 117.7 | 143.8 | 169.8 | 191.9 | 198.2 | 215.5 |  |
|  | SOP | 709 | 11,795 | 13,162 | 15,848 | 7,632 | 2,046 | 1,435 | 481 | 435 | 53,544 |
| 2003 | Numbers | 52.11 | 63.02 | 182.53 | 65.45 | 64.37 | 21.47 | 6.26 | 4.35 | 1.81 | 461.38 |
|  | Mean W. | 13.0 | 37.4 | 76.5 | 113.3 | 132.7 | 142.2 | 153.5 | 169.9 | 162.2 |  |
|  | SOP | 678 | 2,355 | 13,957 | 7,416 | 8,540 | 3,053 | 961 | 740 | 294 | 37,994 |
| 2004 | Numbers | 25.7 | 209.3 | 96.0 | 94.0 | 18.2 | 16.8 | 4.5 | 1.5 | 0.6 | 466.71 |
|  | Mean W. | 27.1 | 43.2 | 81.9 | 117.1 | 145.4 | 157.4 | 170.7 | 184.4 | 187.1 |  |
|  | SOP | 695 | 9,047 | 7,869 | 11,005 | 2,652 | 2,651 | 769 | 279 | 111 | 35,078 |
| 2005 | Numbers | 95.3 | 96.9 | 203.3 | 75.4 | 46.9 | 9.3 | 11.5 | 3.5 | 1.4 | 543.51 |
|  | Mean W. | 14.1 | 54.9 | 85.6 | 121.6 | 148.3 | 162.7 | 176.3 | 178.3 | 200.6 |  |
|  | SOP | 1,341 | 5,319 | 17,415 | 9,163 | 6,961 | 1,519 | 2,028 | 618 | 282 | 44,645 |
| $2006{ }^{1}$ | Numbers | 3.7 | 89.7 | 112.1 | 109.5 | 47.2 | 54.4 | 10.6 | 10.1 | 5.1 | 442.3 |
|  | Mean W. | 21.1 | 39.2 | 83.6 | 113.7 | 143.3 | 175.6 | 198.1 | 210.1 | 220.7 |  |
|  | SOP | 77 | 3,512 | 9,369 | 12,446 | 6,766 | 9,550 | 2,100 | 2,118 | 1,131 | 47,070 |

[^5]${ }^{1} 2000$ tonnes of landings from IIIa are missing. See text section 3.1.2

Table 3.2.14 WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP (t)
by age and quarter from. Western Baltic Spring Spawners
(values from the North Sea, see Table 2.2.1-2.2.5)

|  |  | Division: |  | IV + IIIa + 22-24 |  | Year: |  |  | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarter | W-rings | Division IV |  | Division Illa |  | Sub-division 22-24 |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 0.13 | 24 | 46.34 | 20 | 15.04 | 15 | 61.51 | 18 |
|  | 2 | 0.45 | 107 | 53.38 | 70 | 27.15 | 45 | 80.99 | 62 |
|  | 3 | 1.00 | 127 | 59.38 | 101 | 46.63 | 87 | 107.00 | 95 |
|  | 4 | 0.15 | 140 | 14.38 | 122 | 30.40 | 115 | 44.93 | 117 |
|  | 5 | 0.22 | 151 | 20.84 | 179 | 13.50 | 146 | 34.57 | 166 |
|  | 6 | 0.01 | 193 | 2.32 | 188 | 9.11 | 167 | 11.44 | 171 |
|  | 7 | 0.04 | 164 | 1.28 | 185 | 8.71 | 186 | 10.03 | 186 |
|  | 8+ | 0.03 | 192 | 0.66 | 187 | 4.99 | 187 | 5.68 | 187 |
|  | Total | 2.04 |  | 198.59 |  | 155.53 |  | 356.15 |  |
|  | SOP |  | 249 |  | 16,945 |  | 15,061 |  | 32,254 |
| Quarter |  | Division IV |  | Division IIIa |  | Sub-division 22-24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 0.00 | 0.00 | 0.24 | 41 | 4.88 | 22 | 5.12 | 23 |
|  | 2 | 2.97 | 126 | 6.60 | 76 | 14.68 | 48 | 24.25 | 65 |
|  | 3 | 5.22 | 141 | 4.18 | 105 | 43.78 | 78 | 53.18 | 86 |
|  | 4 | 11.95 | 160 | 1.17 | 124 | 44.91 | 106 | 58.03 | 118 |
|  | 5 | 16.54 | 168 | 1.57 | 161 | 18.96 | 140 | 37.08 | 154 |
|  | 6 | 2.98 | 191 | 0.03 | 179 | 15.64 | 160 | 18.65 | 165 |
|  | 7 | 2.34 | 206 | 0.12 | 179 | 11.93 | 177 | 14.40 | 181 |
|  | 8+ | 1.62 | 222 | 0.15 | 179 | 5.56 | 179 | 7.32 | 188 |
|  | Total | 43.63 |  | 14.05 |  | 160.34 |  | 218.02 |  |
|  | SOP |  | 7,214 |  | 1,397 |  | 17,240 |  | 25,851 |
| Quarter |  | Division IV |  | Division IIIa |  | Sub-division 22-24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.00 | 0.00 | 0.26 | 12 | 0.00 | 22 | 0.26 | 12 |
|  | 1 | 0.00 | 110.50 | 11.96 | 62 | 10.79 | 42 | 22.75 | 52 |
|  | 2 | 0.05 | 148.70 | 34.01 | 97 | 10.20 | 60 | 44.26 | 89 |
|  | 3 | 2.40 | 174 | 27.13 | 124 | 9.22 | 65 | 38.75 | 113 |
|  | 4 | 1.33 | 191 | 12.92 | 142 | 12.81 | 54 | 27.07 | 103 |
|  | 5 | 4.50 | 193 | 6.03 | 160 | 5.49 | 57 | 16.02 | 134 |
|  | 6 | 0.95 | 259 | 2.68 | 175 | 4.17 | 59 | 7.81 | 123 |
|  | 7 | 0.95 | 246 | 2.55 | 188 | 0.73 | 75 | 4.23 | 181 |
|  | 8+ | 0.68 | 255 | 0.94 | 204 | 1.00 | 64 | 2.63 | 164 |
|  | Total | 10.87 |  | 98.48 |  | 54.40 |  | 163.76 |  |
|  | SOP |  | 2,203 |  | 11,356 |  | 3,029 |  | 16,587 |
| Quarter |  | Division IV |  | Division IIIa |  | Sub-division 22-24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 0.00 | 0 | 3.40 | 22 | 0.65 | 21 | 4.05 | 22 |
|  | 1 | 0.00 | 141 | 30.99 | 60 | 14.06 | 53 | 45.06 | 58 |
|  | 2 | 0.04 | 165 | 14.57 | 94 | 20.04 | 77 | 34.65 | 84 |
|  | 3 | 0.16 | 182 | 10.02 | 134 | 19.37 | 99 | 29.56 | 112 |
|  | 4 | 0.52 | 203 | 4.80 | 153 | 13.61 | 106 | 18.92 | 120 |
|  | 5 | 1.11 | 210 | 3.58 | 190 | 5.05 | 89 | 9.74 | 140 |
|  | 6 | 1.16 | 236 | 0.46 | 209 | 2.44 | 102 | 4.07 | 152 |
|  | 7 | 1.92 | 237 | 0.86 | 232 | 0.74 | 143 | 3.53 | 216 |
|  | 8+ | 0.72 | 255 | 0.32 | 208 | 0.61 | 122 | 1.65 | 197 |
|  | Total | 5.64 |  | 69.00 |  | 76.58 |  | 151.22 |  |
|  | SOP |  | 1,288 |  | 6,419 |  | 6,531 |  | 14,238 |
| Quarter |  | Division IV |  | Division IIIa |  | Sub-division 22-24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| T | 0 | 0.00 | 0 | 3.66 | 21 | 0.65 | 21 | 4.31 | 21 |
|  | 1 | 0.13 | 25 | 89.54 | 39 | 44.76 | 34 | 134.43 | 37 |
| 0 | 2 | 3.51 | 125 | 108.56 | 82 | 72.07 | 57 | 184.14 | 73 |
|  | 3 | 8.78 | 149 | 100.71 | 111 | 119.00 | 84 | 228.48 | 98 |
|  | 4 | 13.96 | 164 | 33.26 | 135 | 101.73 | 102 | 148.95 | 115 |
| t | 5 | 22.37 | 175 | 32.03 | 176 | 43.00 | 125 | 97.40 | 153 |
|  | 6 | 5.10 | 214 | 5.50 | 183 | 31.36 | 144 | 41.97 | 158 |
| a | 7 | 5.26 | 224 | 4.82 | 195 | 22.11 | 176 | 32.19 | 187 |
|  | 8+ | 3.06 | 237 | 2.07 | 197 | 12.16 | 170 | 17.28 | 185 |
|  | Total | 62.17 |  | 380.13 |  | 446.84 |  | 889.15 |  |
|  | SOP |  | 10,953 |  | 36,116 |  | 41,861 |  | 88,931 |

Table 3.2.15
WESTERN BALTIC HERRING.
Total catch in numbers (mill) of Western Baltic Spring Spawners in Division IIIe and the North Sea + in Sub-Divisions 22-24 in the years 1991-2006

| Year Area |  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |
| 1991 | Div. IV+Div. IIIa | 100.0 | 157.4 | 382.9 | 394.8 | 167.0 | 112.4 | 21.9 | 7.3 | 3.2 | 1246.8 |
|  | Sub-div. 22-24 | 19.0 | 668.5 | 158.3 | 169.7 | 112.8 | 65.1 | 24.6 | 5.9 | 1.8 | 1206.8 |
| 1992 | Div. IV+Div. IIIa | 109.1 | 246.0 | 321.9 | 174.0 | 154.5 | 78.3 | 55.8 | 17.9 | 8.5 | 1056.9 |
|  | Sub-div. 22-24 | 36.0 | 210.7 | 280.8 | 190.8 | 179.5 | 104.9 | 84.0 | 34.8 | 14.0 | 1099.5 |
| 1993 | Div. IV+Div. IIIa | 161.3 | 371.5 | 315.8 | 219.0 | 94.1 | 59.4 | 41.0 | 21.7 | 8.2 | 1130.8 |
|  | Sub-div. 22-24 | 44.9 | 159.2 | 180.1 | 196.1 | 166.9 | 151.1 | 61.8 | 42.2 | 16.3 | 973.7 |
| 1994 | Div. IV+Div. IIIa | 60.6 | 153.1 | 261.1 | 221.6 | 131.0 | 77.3 | 44.4 | 14.4 | 8.6 | 911.6 |
|  | Sub-div. 22-24 | 202.6 | 96.3 | 103.8 | 161.0 | 136.1 | 90.8 | 74.0 | 35.1 | 24.5 | 721.6 |
| 1995 | Div. IV+Div. IIIa | 50.3 | 302.5 | 204.2 | 97.9 | 90.9 | 30.6 | 21.3 | 12.0 | 7.2 | 816.9 |
|  | Sub-div. 22-24 | 491.0 | 1,358.2 | 233.9 | 128.9 | 104.0 | 53.6 | 38.8 | 20.9 | 13.2 | 1951.5 |
| 1996 | Div. IV+Div. IIIa | 166.2 | 228.1 | 317.7 | 75.6 | 40.4 | 30.6 | 12.6 | 6.7 | 5.6 | 883.6 |
|  | Sub-div. 22-24 | 4.9 | 410.8 | 82.8 | 124.1 | 103.7 | 99.5 | 52.7 | 24.0 | 19.5 | 917.1 |
| 1997 | Div. IV+Div. IIIa | 26.0 | 73.4 | 158.7 | 180.1 | 30.2 | 14.2 | 4.8 | 1.8 | 2.3 | 491.3 |
|  | Sub-div. 22-24 | 350.8 | 595.2 | 130.6 | 96.9 | 45.1 | 29.0 | 35.1 | 19.5 | 21.8 | 973.2 |
| 1998 | Div. IV+Div. IIIa | 36.3 | 175.1 | 315.1 | 94.5 | 54.7 | 11.2 | 8.7 | 2.2 | 2.1 | 700.0 |
|  | Sub-div. 22-24 | 513.5 | 447.9 | 115.8 | 88.3 | 92.0 | 34.1 | 15.0 | 13.2 | 12.0 | 818.4 |
| 1999 | Div. IV+Div. IIIa | 41.34 | 190.29 | 155.67 | 122.26 | 43.16 | 22.21 | 4.42 | 3.02 | 2.40 | 584.8 |
|  | Sub-div. 22-24 | 528.3 | 425.8 | 178.7 | 123.9 | 47.1 | 33.7 | 11.1 | 6.5 | 3.7 | 830.5 |
| 2000 | Div. IV+Div. IIIa | 114.8 | 318.2 | 302.1 | 99.9 | 50.8 | 18.8 | 8.2 | 1.3 | 1.4 | 915.6 |
|  | Sub-div. 22-24 | 37.7 | 616.3 | 194.3 | 86.7 | 77.8 | 53.0 | 30.1 | 12.4 | 9.3 | 1079.9 |
| 2001 | Div. IV+Div. IIIa | 121.7 | 36.6 | 208.1 | 111.1 | 32.1 | 19.7 | 9.8 | 4.2 | 2.4 | 545.6 |
|  | Sub-div. 22-24 | 634.6 | 486.5 | 280.7 | 146.8 | 76.0 | 48.7 | 29.3 | 14.1 | 4.3 | 1721.0 |
| 2002 | Div. IV+Div. IIIa | 69.6 | 577.7 | 168.3 | 134.6 | 53.1 | 12.0 | 7.5 | 2.4 | 2.0 | 1027.3 |
|  | Sub-div. 22-24 | 80.6 | 81.4 | 113.6 | 186.7 | 119.2 | 45.1 | 31.1 | 11.4 | 6.3 | 675.4 |
| 2003 | Div. IV+Div. IIIa | 52.1 | 63.0 | 182.5 | 64.0 | 62.2 | 20.3 | 5.9 | 3.8 | 1.6 | 455.5 |
|  | Sub-div. 22-24 | 1.4 | 63.9 | 82.3 | 95.8 | 125.1 | 82.2 | 22.9 | 13.1 | 7.0 | 493.6 |
| 2004 | Div. IV+Div. IIIa | 25.7 | 209.3 | 96.0 | 94.0 | 18.2 | 16.8 | 4.5 | 1.5 | 0.6 | 466.7 |
|  | Sub-div. 22-24 | 217.9 | 248.4 | 101.8 | 70.8 | 75.0 | 74.4 | 44.5 | 13.4 | 10.4 | 856.5 |
| 2005 | Div. IV+Div. IIIa | 95.3 | 96.9 | 203.3 | 75.4 | 46.9 | 9.3 | 11.5 | 3.5 | 1.4 | 543.5 |
|  | Sub-div. 22-24 | 11.6 | 207.6 | 115.9 | 102.5 | 83.5 | 51.3 | 54.2 | 27.8 | 11.2 | 665.5 |
| $2006{ }^{1}$ | Div. IV+Div. IIIa | 3.7 | 89.7 | 112.1 | 109.5 | 47.2 | 54.4 | 10.6 | 10.1 | 5.1 | 442.3 |
|  | Sub-div. 22-24 | 0.6 | 44.8 | 72.1 | 119.0 | 101.7 | 43.0 | 31.4 | 22.1 | 12.2 | 446.8 |

Data for 1995-2001 for the North Sea and Div. IIIa was revised in 2003.
${ }^{1} 2000$ tonnes of landings from IIIa are missing, and a proportion of those are autumn spawners. See text section 3.1.2

Table 3.2.16 WESTERN BALTIC HERRING.
Mean weight ( $\mathbf{g}$ ) and SOP (tons) of Western Baltic Spring Spawners in Division Illa and the North Sea + in Sub-Divisions 22-24 in the years 1991-2006

| Year Area |  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | SOP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |
| 1991 | Div. IV+Div. IIIa | 33.0 | 48.6 | 69.5 | 99.9 | 135.7 | 146.2 | 166.9 | 179.7 | 193.2 | 121,687 |
|  | Sub-div. 22-24 | 11.5 | 31.5 | 60.4 | 83.2 | 105.2 | 126.6 | 145.6 | 160.0 | 163.7 | 69,886 |
| 1992 | Div. IV+Div. IIIa | 13.9 | 44.1 | 87.0 | 112.9 | 136.2 | 166.3 | 183.5 | 194.4 | 203.6 | 109,523 |
|  | Sub-div. 22-24 | 19.1 | 23.3 | 44.8 | 77.4 | 99.2 | 123.3 | 152.9 | 166.2 | 184.2 | 84,888 |
| 1993 | Div. IV+Div. IIIa | 15.1 | 25.9 | 81.4 | 127.5 | 150.1 | 171.1 | 195.9 | 209.1 | 239.0 | 104,498 |
|  | Sub-div. 22-24 | 16.2 | 24.5 | 44.5 | 73.6 | 94.1 | 122.4 | 149.4 | 168.5 | 178.7 | 80,512 |
| 1994 | Div. IV+Div. IIIa | 20.2 | 42.6 | 94.8 | 122.7 | 150.3 | 168.7 | 194.7 | 209.9 | 220.2 | 106,013 |
|  | Sub-div. 22-24 | 12.9 | 28.2 | 54.2 | 76.4 | 95.0 | 117.7 | 133.6 | 154.3 | 173.9 | 66,425 |
| 1995 | Div. IV+Div. IIIa | 17.9 | 41.5 | 97.8 | 138.0 | 163.1 | 198.5 | 207.0 | 228.8 | 234.3 | 76,674 |
|  | Sub-div. 22-24 | 9.3 | 16.3 | 42.8 | 68.3 | 88.9 | 125.4 | 150.4 | 193.3 | 207.4 | 74,157 |
| 1996 | Div. IV+Div. IIIa | 10.5 | 27.6 | 90.1 | 134.9 | 164.9 | 186.6 | 204.1 | 208.5 | 220.2 | 64,449 |
|  | Sub-div. 22-24 | 12.1 | 22.9 | 45.8 | 74.0 | 92.1 | 116.3 | 120.8 | 139.0 | 182.5 | 56,817 |
| 1997 | Div. IV+Div. IIIa | 19.2 | 49.7 | 76.7 | 127.2 | 154.4 | 175.8 | 184.4 | 192.0 | 208.0 | 48,075 |
|  | Sub-div. 22-24 | 30.4 | 24.7 | 58.4 | 101.0 | 120.7 | 155.2 | 181.3 | 197.1 | 208.8 | 67,513 |
| 1998 | Div. IV+Div. IIIa | 27.8 | 51.3 | 71.5 | 108.8 | 142.6 | 171.7 | 194.4 | 184.2 | 230.0 | 55,121 |
|  | Sub-div. 22-24 | 13.3 | 26.3 | 52.2 | 78.6 | 103.0 | 125.2 | 150.0 | 162.1 | 179.5 | 51,911 |
| 1999 | Div. IV+Div. IIIa | 11.5 | 51.0 | 83.6 | 114.9 | 121.2 | 145.2 | 169.6 | 123.8 | 152.3 | 47,179 |
|  | Sub-div. 22-24 | 11.1 | 26.9 | 50.4 | 81.6 | 112.0 | 148.4 | 151.4 | 167.8 | 161.0 | 50,060 |
| 2000 | Div. IV+Div. IIIa | 22.6 | 31.9 | 67.4 | 107.7 | 140.2 | 170.0 | 157.0 | 185.0 | 210.1 | 56,010 |
|  | Sub-div. 22-24 | 16.5 | 22.2 | 42.8 | 80.4 | 123.5 | 133.2 | 143.4 | 155.4 | 151.4 | 53,904 |
| 2001 | Div. IV+Div. IIIa | 9.0 | 51.2 | 76.2 | 108.9 | 145.3 | 171.4 | 188.2 | 187.2 | 203.3 | 42,079 |
|  | Sub-div. 22-24 | 12.9 | 22.3 | 46.8 | 69.0 | 93.5 | 150.8 | 145.1 | 146.3 | 153.1 | 63,724 |
| 2002 | Div. IV+Div. IIIa | 10.2 | 20.4 | 78.2 | 117.7 | 143.8 | 169.8 | 191.9 | 198.2 | 215.5 | 53,544 |
|  | Sub-div. 22-24 | 10.8 | 27.3 | 57.8 | 81.7 | 108.8 | 132.1 | 186.6 | 177.8 | 157.7 | 52,647 |
| 2003 | Div. IV+Div. IIIa | 13.0 | 37.4 | 76.5 | 112.7 | 132.1 | 140.8 | 151.9 | 167.4 | 158.2 | 37,075 |
|  | Sub-div. 22-24 | 22.4 | 25.8 | 46.4 | 75.3 | 95.2 | 117.2 | 125.9 | 157.1 | 162.6 | 40,315 |
| 2004 | Div. IV+Div. IIIa | 27.1 | 43.2 | 81.9 | 117.1 | 145.4 | 157.4 | 170.7 | 184.4 | 187.1 | 35,078 |
|  | Sub-div. 22-24 | 3.7 | 14.3 | 47.4 | 77.7 | 96.4 | 125.5 | 150.4 | 165.8 | 151.0 | 41,736 |
| 2005 | Div. IV+Div. IIIa | 14.1 | 54.9 | 85.6 | 121.6 | 148.3 | 162.7 | 176.3 | 178.3 | 200.6 | 44,645 |
|  | Sub-div. 22-24 | 13.6 | 14.2 | 48.3 | 73.3 | 89.3 | 115.5 | 143.6 | 159.9 | 170.2 | 43,725 |
| $2006{ }^{1}$ | Div. IV+Div. IIIa | 21.1 | 39.2 | 83.6 | 113.7 | 143.3 | 175.6 | 198.1 | 210.1 | 220.7 | 47,070 |
|  | Sub-div. 22-24 | 21.2 | 34.0 | 56.7 | 84.0 | 102.2 | 125.3 | 143.9 | 175.8 | 170.0 | 41,861 |

Data for 1995-2001 for the North Sea and Div. IIIa was revised in 2003.
${ }^{1} 2000$ tonnes of landings from IIIa are missing, and a proportion of those are autumn spawners. See text section 3.1.2

Table 3.2.17 WESTERN BALTIC HERRING.
Transfers of North Sea autumn spawners from Div. Illa to the North Sea Numbers (mill) and mean weight, SOP in (tonnes) 1991-2006.

|  | W-Rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year |  |  |  |  |  |  |  |  |  |  |  |
| 1991 | Number | 677.1 | 748.3 | 298.3 | 52.4 | 7.7 | 5.1 | 1.1 | 0.4 | 0.1 | 1,790.6 |
|  | Mean W. | 25.6 | 40.5 | 72.9 | 97.2 | 135.8 | 149.7 | 155.7 | 159.8 | 176.8 |  |
|  | SOP | 17,314 | 30,336 | 21,744 | 5,098 | 1,049 | 771 | 178 | 59 | 26 | 76,575 |
| 1992 | Number | 2,298.4 | 1,408.8 | 220.3 | 22.1 | 10.4 | 6.6 | 2.9 | 1.0 | 0.4 | 3,970.9 |
|  | Mean W. | 12.3 | 51.8 | 84.2 | 131.4 | 162.0 | 173.4 | 185.3 | 198.4 | 201.2 |  |
|  | SOP | 28,159 | 72,985 | 18,557 | 2,907 | 1,683 | 1,143 | 533 | 200 | 84 | 126,251 |
| 1993 | Number | 2,795.4 | 2,032.5 | 237.6 | 26.5 | 7.7 | 3.6 | 2.7 | 2.2 | 0.7 | 5,109.0 |
|  | Mean W. | 12.5 | 28.6 | 79.7 | 141.4 | 132.3 | 233.4 | 238.5 | 180.6 | 203.1 |  |
|  | SOP | 34,903 | 58,107 | 18,939 | 3,749 | 1,016 | 850 | 647 | 390 | 133 | 118,734 |
| 1994 | Number | 481.6 | 1,086.5 | 201.4 | 26.9 | 6.0 | 2.9 | 1.6 | 0.4 | 0.2 | 1,807.5 |
|  | Mean W. | 16.0 | 42.9 | 83.4 | 110.7 | 138.3 | 158.6 | 184.6 | 199.1 | 213.9 |  |
|  | SOP | 7,723 | 46,630 | 16,790 | 2,980 | 831 | 460 | 287 | 75 | 37 | 75,811 |
| 1995 | Number | 1,144.5 | 1,189.2 | 161.5 | 13.3 | 3.5 | 1.1 | 0.6 | 0.4 | 0.3 | 2,514.4 |
|  | Mean W. | 11.2 | 39.1 | 88.3 | 145.7 | 165.5 | 204.5 | 212.2 | 236.4 | 244.3 |  |
|  | SOP | 12,837 | 46,555 | 14,267 | 1,940 | 573 | 225 | 133 | 86 | 65 | 76,680 |
| 1996 | Number | 516.1 | 961.1 | 161.4 | 17.0 | 3.4 | 1.6 | 0.7 | 0.4 | 0.3 | 1,661.9 |
|  | Mean W. | 11.0 | 23.4 | 80.2 | 126.6 | 165.0 | 186.5 | 216.1 | 216.3 | 239.1 |  |
|  | SOP | 5,697 | 22,448 | 12,947 | 2,151 | 565 | 307 | 145 | 77 | 66 | 44,403 |
| 1997 | Number | 67.6 | 305.3 | 131.7 | 21.2 | 1.7 | 0.8 | 0.2 | 0.1 | 0.1 | 528.7 |
|  | Mean W. | 19.3 | 47.7 | 68.5 | 124.4 | 171.5 | 184.7 | 188.7 | 188.7 | 192.4 |  |
|  | SOP | 1,304 | 14,571 | 9,025 | 2,643 | 285 | 146 | 40 | 16 | 25 | 28,057 |
| 1998 | Number | 51.3 | 745.1 | 161.5 | 26.6 | 19.2 | 3.0 | 3.1 | 1.2 | 0.5 | 1,011.6 |
|  | Mean W. | 27.4 | 56.4 | 79.8 | 117.8 | 162.9 | 179.7 | 197.2 | 178.9 | 226.3 |  |
|  | SOP | 1,409 | 41,994 | 12,896 | 3,137 | 3,136 | 547 | 608 | 211 | 108 | 64,045 |
| 1999 | Number | 598.8 | 303.0 | 148.6 | 47.2 | 13.4 | 6.2 | 1.2 | 0.5 | 0.5 | 1,119.4 |
|  | Mean W. | 10.4 | 50.5 | 87.7 | 113.7 | 137.4 | 156.5 | 188.1 | 187.3 | 198.8 |  |
|  | SOP | 6,255 | 15,297 | 13,037 | 5,369 | 1,841 | 974 | 230 | 90 | 92 | 43,186 |
| 2000 | Number | 235.3 | 984.3 | 116.0 | 21.9 | 22.9 | 7.5 | 3.3 | 0.6 | 0.1 | 1,391.8 |
|  | Mean W. | 21.3 | 28.5 | 76.1 | 108.8 | 163.1 | 190.3 | 183.9 | 189.4 | 200.2 |  |
|  | SOP | 5,005 | 28,012 | 8,825 | 2,377 | 3,731 | 1,436 | 601 | 114 | 13 | 50,115 |
| 2001 | Number | 807.8 | 563.6 | 150.0 | 17.2 | 1.4 | 0.3 | 0.5 | 0.0 | 0.0 | 1,540.8 |
|  | Mean W. | 8.7 | 49.4 | 75.3 | 108.2 | 130.1 | 147.1 | 219.1 | 175.8 | 198.1 |  |
|  | SOP | 7,029 | 27,849 | 11,300 | 1,856 | 177 | 43 | 109 | 8 | 5 | 48,376 |
| 2002 | Number | 478.5 | 362.6 | 56.7 | 5.6 | 0.7 | 0.2 | 0.1 | 0.0 | 0.0 | 904.5 |
|  | Mean W. | 12.2 | 38.0 | 100.6 | 121.5 | 142.7 | 160.9 | 178.7 | 177.4 | 218.6 |  |
|  | SOP | 5,859 | 13,790 | 5,705 | 684 | 106 | 26 | 21 | 8 | 5 | 26,205 |
| 2003 | Number | 21.6 | 445.0 | 182.3 | 13.0 | 16.2 | 1.8 | 1.1 | 1.2 | 0.2 | 682.4 |
|  | Mean W. | 20.5 | 33.7 | 67.0 | 123.2 | 150.3 | 163.5 | 190.2 | 214.6 | 186.8 |  |
|  | SOP | 442 | 14,992 | 12,219 | 1,606 | 2,436 | 293 | 213 | 264 | 33 | 32,498 |
| 2004 | Number | 88.4 | 70.9 | 179.9 | 20.7 | 6.0 | 9.7 | 1.8 | 2.0 | 0.9 | 380.4 |
|  | Mean W. | 22.5 | 55.3 | 70.2 | 120.6 | 140.9 | 151.7 | 170.6 | 186.6 | 178.5 |  |
|  | SOP | 1,993 | 3,921 | 12,638 | 2,498 | 851 | 1,479 | 312 | 367 | 154 | 24,214 |
| 2005 | Number | 96.4 | 307.5 | 159.2 | 16.2 | 5.4 | 2.4 | 2.3 | 0.5 | 0.2 | 589.9 |
|  | Mean W. | 16.5 | 50.5 | 71.0 | 105.9 | 154.6 | 173.5 | 184.5 | 200.2 | 208.9 |  |
|  | SOP | 1,595 | 15,527 | 11,304 | 1,712 | 828 | 412 | 420 | 95 | 34 | 31,927 |
| $2006{ }^{1}$ | Number | 35.1 | 150.1 | 50.2 | 10.2 | 3.3 | 3.3 | 0.6 | 0.4 | 0.2 | 253.3 |
|  | Mean W. | 14.3 | 53.5 | 79.2 | 117.6 | 140.2 | 185.5 | 190.4 | 215.6 | 206.9 |  |
|  | SOP | 503 | 8,035 | 3,975 | 1,200 | 456 | 620 | 107 | 81 | 37 | 15,015 |

[^6]${ }^{1} 2000$ tonnes of landings from IIIa are missing, and a proportion of those are autumn spawners. See text section 3.1.2

Table 3.3.1 WESTERN BALTIC HERRING.
International Bottom Trawl Survey (IBTS) in the Kattegat in quarter 1.
Mean catch of spring-spawning herring at age in number per hour.

| Year | Winter rings |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ |
| $\mathbf{1 9 9 0}$ | 416 | 681 | 65 | 43 | 11 |
| $\mathbf{1 9 9 1}$ | 190 | 206 | 144 | 25 | 20 |
| $\mathbf{1 9 9 2}$ | 588 | 82 | 33 | 21 | 13 |
| $\mathbf{1 9 9 3}$ | 3140 | 554 | 81 | 35 | 50 |
| $\mathbf{1 9 9 4}$ | 1380 | 256 | 112 | 22 | 31 |
| $\mathbf{1 9 9 5}$ | 781 | 132 | 30 | 42 | 24 |
| $\mathbf{1 9 9 6}$ | 1312 | 1405 | 160 | 42 | 22 |
| $\mathbf{1 9 9 7}$ | 3267 | 229 | 119 | 15 | 18 |
| $\mathbf{1 9 9 8}$ | 407 | 853 | 165 | 74 | 8 |
| $\mathbf{1 9 9 9}$ | 309 | 66 | 43 | 21 | 14 |
| $\mathbf{2 0 0 0}$ | 1933 | 219 | 28 | 10 | 7 |
| $\mathbf{2 0 0 1 *}$ | - | - | - | - | - |
| $\mathbf{2 0 0 2}$ | 2335 | 178 | 222 | 23 | 7 |
| $\mathbf{2 0 0 3}$ | 1364 | 1495 | 41 | 10 | 0 |
| $\mathbf{2 0 0 4}$ | 147 | 144 | 37 | 6 | 2 |
| $\mathbf{2 0 0 5}$ | 286 | 257 | 26 | 12 | 5 |
| $\mathbf{2 0 0 6}$ | 361 | 163 | 48 | 19 | 17 |
| $\mathbf{2 0 0 7}$ | 346 | 185 | 15 | 10 | 0 |
| * no data available |  |  |  |  |  |

Table 3.3.2 WESTERN BALTIC HERRING.
International Bottom Trawl Survey (IBTS) in the Kattegat in quarter 3.
Mean catch of spring-spawning herring at age in number per hour.

| Year | Winter rings |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ |
| $\mathbf{1 9 9 1}$ | 141 | 83 | 101 | 41 | 24 |
| $\mathbf{1 9 9 2}$ | 372 | 108 | 70 | 63 | 25 |
| $\mathbf{1 9 9 3}$ | 404 | 159 | 42 | 36 | 25 |
| $\mathbf{1 9 9 4}$ | 265 | 229 | 154 | 49 | 36 |
| $\mathbf{1 9 9 5}$ | 687 | 192 | 113 | 99 | 29 |
| $\mathbf{1 9 9 6}$ | 631 | 322 | 31 | 17 | 11 |
| $\mathbf{1 9 9 7}$ | 52 | 122 | 33 | 8 | 13 |
| $\mathbf{1 9 9 8}$ | 118 | 86 | 22 | 27 | 5 |
| $\mathbf{1 9 9 9}$ | 292 | 116 | 71 | 34 | 14 |
| $\mathbf{2 0 0 0}$ | - | - | - | - | - |
| $\mathbf{2 0 0 1}$ | 313 | 190 | 72 | 18 | 2 |
| $\mathbf{2 0 0 2}$ | 1568 | 169 | 100 | 16 | 6 |
| $\mathbf{2 0 0 3}$ | 969 | 550 | 170 | 53 | 29 |
| $\mathbf{2 0 0 4}$ | 1225 | 215 | 144 | 30 | 23 |
| $\mathbf{2 0 0 5}$ | 607 | 255 | 54 | 23 | 13 |
| $\mathbf{2 0 0 6}$ | 509 | 79 | 64 | 40 | 32 |
| no survey was carried out in 2000 |  |  |  |  |  |

Table 3.3.3 WESTERN BALTIC HERRING. Acoustic surveys on the Spring Spawning Herring in the North Sea/Division IIIa in 1991-2006 (July).


| Biomass ('000 tonnnes) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 |  | 34.3 | 1 | 8.7 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 26.8 | 7 | 0.4 | 77.4 | 52.9 | 4.7 | 7.1 | 74.8 | 61.4 | 3.5 | 137.2 | 79.0 | 63.9 | 105.9 | 112.6 |
| 2 | 177.1 | 169.0 | 139 | 33.2 | 108.9 | 87.0 | 52.2 | 136.1 | 101.6 | 138.1 | 55.8 | 107.2 | 91.5 | 75.6 | 100.1 | 160.5 |
| 3 | 219.7 | 206.3 | 112 | 114.7 | 102.6 | 27.6 | 81.0 | 84.8 | 59.5 | 68.8 | 51.2 | 126.9 | 41.4 | 89.4 | 46.6 | 158.6 |
| 4 | 116.0 | 204.7 | 69 | 76.7 | 145.5 | 17.9 | 21.5 | 35.2 | 14.7 | 45.3 | 21.5 | 55.9 | 41.7 | 41.5 | 28.9 | 56.3 |
| 5 | 51.1 | 83.3 | 65 | 41.8 | 33.9 | 17.8 | 9.8 | 13.1 | 3.4 | 25.1 | 17.9 | 12.8 | 13.9 | 29.3 | 16.5 | 23.7 |
| 6 | 19.0 | 36.6 | 26 | 38.1 | 27.4 | 5.8 | 9.8 | 6.9 | 0.5 | 10.0 | 6.9 | 7.4 | 4.2 | 11.7 | 14.9 | 4.1 |
| 7 | 13.0 | 24.4 | 16 | 13.1 | 6.7 | 3.3 | 14.9 | 4.8 | 0.3 | 1.4 | 4.7 | 3.5 | 2.0 | 4.1 | 7.5 | 1.6 |
| 8+ | 2.0 | 5.0 | 2 | 7.8 | 3.8 | 2.7 | 13.6 | 9.0 | 0.1 | 1.3 | 2.7 | 3.1 | 0.9 | 3.2 | 4.9 | 0.02 |
| Total | 597.9 | 756.1 | 436.5 | 325.8 | 506.2 | 215.1 | 207.5 | 297.0 | 254.9 | 351.4 | 164.2 | 454.0 | 274.5 | 318.8 | 325.3 | 517.5 |
| 3+ group | 420.9 | 560.3 | 291.0 | 292.3 | 319.9 | 75.2 | 150.6 | 153.7 | 78.5 | 151.9 | 104.9 | 209.6 | 104.0 | 179.3 | 119.3 | 244.4 |

## Mean weight (g)

| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  | 8.9 | 4.0 | 9.0 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 96.8 | 66.3 | 80.0 | 35.2 | 48.5 | 36.9 | 51.9 | 54.7 | 40.7 | 54.0 | 41.0 | 43.1 | 38.3 | 39.4 | 54.1 |
| 2 | 95.0 | 80.8 | 50.1 | 80.3 | 57.7 | 86.6 | 73.0 | 80.9 | 88.9 | 73.1 | 87.0 | 68.0 | 82.5 | 81.3 | 74.6 | 72.4 |
| 3 | 114.0 | 114.7 | 87.9 | 122.7 | 100.4 | 111.9 | 103.0 | 94.1 | 113.8 | 102.2 | 113.2 | 91.1 | 104.9 | 123.2 | 100.5 | 89.1 |
| 4 | 134.0 | 128.5 | 116.2 | 153.0 | 114.6 | 126.8 | 129.6 | 124.7 | 109.1 | 124.4 | 140.5 | 106.6 | 128.8 | 135.2 | 143.7 | 114.8 |
| 5 | 146.0 | 149.8 | 149.9 | 175.1 | 132.9 | 149.4 | 145.0 | 118.7 | 120.0 | 135.4 | 185.2 | 145.8 | 134.2 | 159.4 | 160.9 | 131.6 |
| 6 | 216.0 | 185.7 | 169.6 | 205.0 | 157.2 | 157.3 | 143.1 | 135.8 | 179.9 | 179.2 | 182.6 | 186.5 | 165.4 | 162.9 | 177.7 | 153.2 |
| 7 | 181.0 | 199.7 | 256.9 | 212.0 | 172.9 | 166.8 | 185.6 | 156.4 | 179.9 | 208.8 | 206.3 | 198.7 | 167.2 | 191.6 | 202.3 | 169.2 |
| 8+ | 200.0 | 252.0 | 164.2 | 230.3 | 183.1 | 212.9 | 178.0 | 168.0 | 181.7 | 135.2 | 226.9 | 183.4 | 170.3 | 178.0 | 229.2 | 178.0 |
| Total | 115.6 | 123.9 | 75.8 | 100.2 | 73.7 | 80.5 | 99.4 | 91.4 | 78.5 | 74.8 | 110.9 | 64.8 | 72.1 | 81.2 | 65.9 | 76.3 |

* revised in 1997
**the survey only covered the Skagerrak area by Norway. Additional estimates for the Kattegat area were added (see ICES 2000/ACFM:10, Table 3.5.8)

Table 3.3.4 WESTERN BALTIC HERRING. Acoustic survey on the Spring Spawning Herring in Sub-divisions 22-24 in 1991-2006 (September/October).

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Numbers in millions |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 5,577 | 3,467 | 768 | 4,383 | 4,001 | 1,418 | 2,608 | 2,179 | 4,821 | 1,021 | 1,831 | 3,984 | 3,701 | 2,401 | 2,769 | 3,438 |
| 1 | 2,507 | 2,179 | 345 | 412 | 1,163 | 1,084 | 1,389 | 451 | 1,145 | 1,208 | 1,314 | 611 | 781 | 912 | 662 | 716 |
| 2 | 880 | 1,015 | 354 | 823 | 307 | 541 | 492 | 557 | 246 | 477 | 1,761 | 372 | 200 | 590 | 569 | 201 |
| 3 | 852 | 465 | 485 | 540 | 332 | 413 | 343 | 364 | 187 | 348 | 1,013 | 566 | 230 | 352 | 378 | 328 |
| 4 | 259 | 233 | 381 | 433 | 342 | 282 | 151 | 232 | 129 | 206 | 357 | 337 | 276 | 166 | 183 | 340 |
| 5 | 102 | 71 | 121 | 182 | 247 | 283 | 112 | 99 | 44 | 81 | 92 | 61 | 103 | 145 | 102 | 180 |
| 6 | 49 | 32 | 52 | 56 | 124 | 110 | 92 | 51 | 8 | 39 | 55 | 23 | 41 | 81 | 87 | 130 |
| 7 | 6 | 8 | 28 | 22 | 40 | 44 | 32 | 23 | 1 | 5 | 5 | 3 | 9 | 23 | 25 | 85 |
| 8+ | 27 | 9 | 13 | 2 | 27 | 18 | 46 | 9 | 2 | 4 | 0 | 13 | 11 | 12 | 16 | 30 |
| Total | 10,259 | 7,480 | 2,547 | 6,854 | 6,583 | 4,193 | 5,265 | 3,966 | 6,582 | 3,389 | 6,428 | 5,970 | 5,353 | 4,682 | 4,791 | 5,447 |
| + group | ,295 | 818 | 1,080 | 1,235 | 1,112 | 1,151 | 775 | 778 | 370 | 682 | 1,522 | 1,002 | 671 | 780 | 79 | ,092 |


| Biomass ('000 tonnnes) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 62.0 | 48.9 | 11.1 | 49.3 | 41.1 | 12.3 | 25.6 | 20.4 | 54.2 | 12.8 | 21.4 | 33.9 | 31.5 | 20.5 | 28.6 | 34.6 |
| 1 | 97.8 | 77.8 | 12.3 | 14.3 | 39.6 | 32.9 | 49.4 | 18.2 | 42.3 | 47.5 | 59.1 | 23.9 | 24.7 | 34.2 | 26.0 | 26.5 |
| 2 | 60.0 | 57.5 | 15.7 | 38.1 | 19.8 | 26.8 | 29.2 | 41.4 | 18.8 | 29.7 | 118.7 | 27.1 | 14.9 | 34.9 | 43.6 | 14.3 |
| 3 | 76.9 | 39.5 | 29.7 | 39.2 | 28.5 | 29.2 | 31.9 | 32.9 | 22.0 | 29.0 | 93.4 | 56.1 | 23.3 | 28.4 | 34.3 | 31.7 |
| 4 | 29.4 | 28.5 | 23.5 | 41.3 | 39.1 | 20.0 | 21.0 | 27.5 | 13.1 | 24.1 | 34.2 | 39.8 | 36.3 | 18.9 | 21.8 | 36.2 |
| 5 | 13.5 | 10.6 | 12.3 | 22.9 | 26.7 | 33.9 | 16.0 | 11.2 | 5.6 | 9.2 | 11.6 | 8.6 | 15.6 | 17.8 | 14.0 | 26.9 |
| 6 | 6.4 | 5.1 | 6.7 | 11.5 | 14.7 | 14.7 | 13.2 | 6.1 | 0.8 | 5.6 | 7.6 | 3.3 | 6.2 | 12.6 | 14.0 | 19.8 |
| 7 | 0.8 | 1.6 | 2.2 | 4.9 | 8.8 | 5.7 | 5.1 | 3.7 | 0.2 | 1.1 | 0.9 | 0.5 | 1.5 | 3.5 | 5.0 | 14.6 |
| 8+ | 3.6 | 2.1 | 1.8 | 0.6 | 6.6 | 2.7 | 10.2 | 2.2 | 0.4 | 0.7 | 0.0 | 1.9 | 1.8 | 2.1 | 3.5 | 6.5 |
| Total | 350.3 | 271.6 | 115.3 | 222.1 | 224.8 | 178.4 | 201.6 | 163.5 | 157.4 | 159.7 | 346.9 | 195.2 | 155.8 | 172.8 | 190.8 | 211.2 |
| 3+ group | 130.5 | 87.4 | 76.2 | 120.4 | 124.4 | 106.3 | 97.4 | 83.5 | 42.1 | 69.6 | 147.7 | 110.3 | 84.6 | 83.2 | 92.6 | 135.7 |


| Mean weight (g) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 11.1 | 14.1 | 14.4 | 11.2 | 10.3 | 8.7 | 9.8 | 9.4 | 11.2 | 12.6 | 11.7 | 8.5 | 8.5 | 8.6 | 10.3 | 10.1 |
| 1 | 39.0 | 35.7 | 35.7 | 34.7 | 34.0 | 30.4 | 35.6 | 40.3 | 37.0 | 39.3 | 45.0 | 39.1 | 31.7 | 37.5 | 39.2 | 37.1 |
| 2 | 68.2 | 56.7 | 44.3 | 46.3 | 64.5 | 49.6 | 59.4 | 74.3 | 76.4 | 62.2 | 67.4 | 72.8 | 74.5 | 59.1 | 76.7 | 71.1 |
| 3 | 90.2 | 84.9 | 61.3 | 72.6 | 85.9 | 70.7 | 93.1 | 90.4 | 117.6 | 83.3 | 92.2 | 99.2 | 101.2 | 80.7 | 90.8 | 96.7 |
| 4 | 113.5 | 122.3 | 61.6 | 95.5 | 114.5 | 71.1 | 139.2 | 118.3 | 101.8 | 117.1 | 95.7 | 118.2 | 131.2 | 113.6 | 118.8 | 106.5 |
| 5 | 132.2 | 148.7 | 101.3 | 125.9 | 108.0 | 119.7 | 142.3 | 114.0 | 127.5 | 114.1 | 126.0 | 142.6 | 151.0 | 122.6 | 137.2 | 149.7 |
| 6 | 130.4 | 161.0 | 129.6 | 204.0 | 118.1 | 133.5 | 143.4 | 120.5 | 107.2 | 143.0 | 137.0 | 142.8 | 150.9 | 154.6 | 161.8 | 153.0 |
| 7 | 133.0 | 205.7 | 80.2 | 222.6 | 222.0 | 128.5 | 161.6 | 158.1 | 232.7 | 202.9 | 175.7 | 205.5 | 155.7 | 151.1 | 202.5 | 171.8 |
| $8+$ | 132.5 | 224.4 | 137.5 | 269.1 | 241.1 | 154.7 | 222.2 | 232.9 | 219.1 | 180.9 |  | 143.5 | 165.6 | 169.0 | 215.3 | 214.1 |
| Total | 34.1 | 36.3 | 45.3 | 32.4 | 34.2 | 42.5 | 38.3 | 41.2 | 23.9 | 47.1 | 54.0 | 32.7 | 29.1 | 36.9 | 39.8 | 38.8 |

${ }^{1)}$ revised in 2001 due to new presented area of strata in the 'Manual for the Baltic
International Acoustic Survey'. ICES CM 2000/H:2 Ref.: D: Annex 3 (Table 2.2)
${ }^{2)}$ incl. estimates for Sub-division 23, which was covered by RV ARGOS (Sweden) in November 2001
${ }^{3)}$ revised in 2003 due to revised Sa values

Table 3.3.4 WESTERN BALTIC HERRING. Acoustic survey on the Spring Spawning Herring in Sub-divisions 22-24 in 1991-2006 (September/October).

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Numbers in millions |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 5,577 | 3,467 | 768 | 4,383 | 4,001 | 1,418 | 2,608 | 2,179 | 4,821 | 1,021 | 1,831 | 3,984 | 3,701 | 2,401 | 2,769 | 3,438 |
| 1 | 2,507 | 2,179 | 345 | 412 | 1,163 | 1,084 | 1,389 | 451 | 1,145 | 1,208 | 1,314 | 611 | 781 | 912 | 662 | 716 |
| 2 | 880 | 1,015 | 354 | 823 | 307 | 541 | 492 | 557 | 246 | 477 | 1,761 | 372 | 200 | 590 | 569 | 201 |
| 3 | 852 | 465 | 485 | 540 | 332 | 413 | 343 | 364 | 187 | 348 | 1,013 | 566 | 230 | 352 | 378 | 328 |
| 4 | 259 | 233 | 381 | 433 | 342 | 282 | 151 | 232 | 129 | 206 | 357 | 337 | 276 | 166 | 183 | 340 |
| 5 | 102 | 71 | 121 | 182 | 247 | 283 | 112 | 99 | 44 | 81 | 92 | 61 | 103 | 145 | 102 | 180 |
| 6 | 49 | 32 | 52 | 56 | 124 | 110 | 92 | 51 | 8 | 39 | 55 | 23 | 41 | 81 | 87 | 130 |
| 7 | 6 | 8 | 28 | 22 | 40 | 44 | 32 | 23 | 1 | 5 | 5 | 3 | 9 | 23 | 25 | 85 |
| $8+$ | 27 | 9 | 13 | 2 | 27 | 18 | 46 | 9 | 2 | 4 | 0 | 13 | 11 | 12 | 16 | 30 |
| Total | 10,259 | 7,480 | 2,547 | 6,854 | 6,583 | 4,193 | 5,265 | 3,966 | 6,582 | 3,389 | 6,428 | 5,970 | 5,353 | 4,682 | 4,791 | 5,447 |
| 3+ group | 1,295 | 818 | 1,080 | 1,235 | 1,112 | 1,151 | 775 | 778 | 370 | 682 | 1,522 | 1,002 | 671 | 780 | 791 | 1,092 |


| Biomass ('000 tonnnes) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 62.0 | 48.9 | 11.1 | 49.3 | 41.1 | 12.3 | 25.6 | 20.4 | 54.2 | 12.8 | 21.4 | 33.9 | 31.5 | 20.5 | 28.6 | 34.6 |
| 1 | 97.8 | 77.8 | 12.3 | 14.3 | 39.6 | 32.9 | 49.4 | 18.2 | 42.3 | 47.5 | 59.1 | 23.9 | 24.7 | 34.2 | 26.0 | 26.5 |
| 2 | 60.0 | 57.5 | 15.7 | 38.1 | 19.8 | 26.8 | 29.2 | 41.4 | 18.8 | 29.7 | 118.7 | 27.1 | 14.9 | 34.9 | 43.6 | 14.3 |
| 3 | 76.9 | 39.5 | 29.7 | 39.2 | 28.5 | 29.2 | 31.9 | 32.9 | 22.0 | 29.0 | 93.4 | 56.1 | 23.3 | 28.4 | 34.3 | 31.7 |
| 4 | 29.4 | 28.5 | 23.5 | 41.3 | 39.1 | 20.0 | 21.0 | 27.5 | 13.1 | 24.1 | 34.2 | 39.8 | 36.3 | 18.9 | 21.8 | 36.2 |
| 5 | 13.5 | 10.6 | 12.3 | 22.9 | 26.7 | 33.9 | 16.0 | 11.2 | 5.6 | 9.2 | 11.6 | 8.6 | 15.6 | 17.8 | 14.0 | 26.9 |
| 6 | 6.4 | 5.1 | 6.7 | 11.5 | 14.7 | 14.7 | 13.2 | 6.1 | 0.8 | 5.6 | 7.6 | 3.3 | 6.2 | 12.6 | 14.0 | 19.8 |
| 7 | 0.8 | 1.6 | 2.2 | 4.9 | 8.8 | 5.7 | 5.1 | 3.7 | 0.2 | 1.1 | 0.9 | 0.5 | 1.5 | 3.5 | 5.0 | 14.6 |
| $8+$ | 3.6 | 2.1 | 1.8 | 0.6 | 6.6 | 2.7 | 10.2 | 2.2 | 0.4 | 0.7 | 0.0 | 1.9 | 1.8 | 2.1 | 3.5 | 6.5 |
| Total | 350.3 | 271.6 | 115.3 | 222.1 | 224.8 | 178.4 | 201.6 | 163.5 | 157.4 | 159.7 | 346.9 | 195.2 | 155.8 | 172.8 | 190.8 | 211.2 |
| 3+ group | 130.5 | 87.4 | 76.2 | 120.4 | 124.4 | 106.3 | 97.4 | 83.5 | 42.1 | 69.6 | 147.7 | 110.3 | 84.6 | 83.2 | 92.6 | 135.7 |


| Mean weight (g) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 11.1 | 14.1 | 14.4 | 11.2 | 10.3 | 8.7 | 9.8 | 9.4 | 11.2 | 12.6 | 11.7 | 8.5 | 8.5 | 8.6 | 10.3 | 10.1 |
| 1 | 39.0 | 35.7 | 35.7 | 34.7 | 34.0 | 30.4 | 35.6 | 40.3 | 37.0 | 39.3 | 45.0 | 39.1 | 31.7 | 37.5 | 39.2 | 37.1 |
| 2 | 68.2 | 56.7 | 44.3 | 46.3 | 64.5 | 49.6 | 59.4 | 74.3 | 76.4 | 62.2 | 67.4 | 72.8 | 74.5 | 59.1 | 76.7 | 71.1 |
| 3 | 90.2 | 84.9 | 61.3 | 72.6 | 85.9 | 70.7 | 93.1 | 90.4 | 117.6 | 83.3 | 92.2 | 99.2 | 101.2 | 80.7 | 90.8 | 96.7 |
| 4 | 113.5 | 122.3 | 61.6 | 95.5 | 114.5 | 71.1 | 139.2 | 118.3 | 101.8 | 117.1 | 95.7 | 118.2 | 131.2 | 113.6 | 118.8 | 106.5 |
| 5 | 132.2 | 148.7 | 101.3 | 125.9 | 108.0 | 119.7 | 142.3 | 114.0 | 127.5 | 114.1 | 126.0 | 142.6 | 151.0 | 122.6 | 137.2 | 149.7 |
| 6 | 130.4 | 161.0 | 129.6 | 204.0 | 118.1 | 133.5 | 143.4 | 120.5 | 107.2 | 143.0 | 137.0 | 142.8 | 150.9 | 154.6 | 161.8 | 153.0 |
| 7 | 133.0 | 205.7 | 80.2 | 222.6 | 222.0 | 128.5 | 161.6 | 158.1 | 232.7 | 202.9 | 175.7 | 205.5 | 155.7 | 151.1 | 202.5 | 171.8 |
| 8+ | 132.5 | 224.4 | 137.5 | 269.1 | 241.1 | 154.7 | 222.2 | 232.9 | 219.1 | 180.9 |  | 143.5 | 165.6 | 169.0 | 215.3 | 214.1 |
| Total | 34.1 | 36.3 | 45.3 | 32.4 | 34.2 | 42.5 | 38.3 | 41.2 | 23.9 | 47.1 | 54.0 | 32.7 | 29.1 | 36.9 | 39.8 | 38.8 |

${ }^{1)}$ revised in 2001 due to new presented area of strata in the 'Manual for the Baltic
International Acoustic Survey'. ICES CM 2000/H:2 Ref.: D: Annex 3 (Table 2.2)
${ }^{2)}$ incl. estimates for Sub-division 23, which was covered by RV ARGOS (Sweden) in November 2001
${ }^{3)}$ revised in 2003 due to revised Sa values

Table 3.3.5 WESTERN BALTIC HERRING. Estimation of the herring 0-Group (TL >=30 mm ) Greifswalder Bodden and adjacent waters (March/April to June).

| Year | Number in Millions |
| :---: | :---: |
| 1977 | $2000^{1}$ |
| 1978 | $100^{1}$ |
| 1979 | $2200{ }^{1}$ |
| 1980 | $360{ }^{1}$ |
| 1981 | $200^{1}$ |
| 1982 | $180^{1}$ |
| 1983 | $1760^{1}$ |
| 1984 | $290{ }^{1}$ |
| 1985 | $1670^{1}$ |
| 1986 | $1500^{1}$ |
| 1987 | $1370^{1}$ |
| 1988 | $1223{ }^{2}$ |
| 1989 | $63^{2}$ |
| 1990 | $57^{2}$ |
| 1991 | $236{ }^{3}$ |
| 1992 | $18^{4}$ |
| 1993 | $199{ }^{4}$ |
| 1994 | $788^{4}$ |
| 1995 | $171{ }^{4}$ |
| 1996 | $31^{4}$ |
| 1997 | $54^{4}$ |
| 1998 | $2553{ }^{4}$ |
| 1999 | $1945{ }^{4}$ |
| 2000 | $151{ }^{4}$ |
| 2001 | $421{ }^{4}$ |
| 2002 | $2051{ }^{4}$ |
| 2003 | $2005{ }^{4}$ |
| 2004 | $860^{4}$ |
| 2005 | $162^{5}$ |
| 2006 | not available |
| ${ }^{1}$ Brieln | 1989 |
| ${ }^{2}$ Klenz 1999 Inf. Fischwirtsch. Fischereiforsch. 46(2), 1999: 15-17 <br> ${ }^{3}$ Müller \& Klenz 1994 <br> ${ }^{4}$ Klenz 2005 Inf. Fischwirtsch. Fischereiforsch. 52, 2005: 21-22 ${ }^{5}$ unpublished |  |
|  |  |
|  |  |

Table 3.6.1 WESTERN BALTIC HERRING. Input to ICA.
Catch in number (millions)

| Age | 1991 | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 119.0 | 145.1 | 206.1 | 263.2 | 541.3 | 171.1 | 376.8 | 549.8 | 569.6 | 152.6 | 756.3 |
| $\mathbf{1}$ | 826.0 | 456.7 | 530.7 | 249.4 | 1660.7 | 638.9 | 668.6 | 623.1 | 616.1 | 934.5 | 523.2 |
| $\mathbf{2}$ | 541.2 | 602.6 | 495.9 | 365.0 | 438.1 | 400.6 | 289.3 | 430.9 | 334.3 | 496.4 | 488.8 |
| $\mathbf{3}$ | 564.4 | 364.9 | 415.1 | 382.6 | 226.8 | 199.7 | 276.9 | 182.9 | 246.2 | 186.6 | 257.8 |
| $\mathbf{4}$ | 279.8 | 334.0 | 260.9 | 267.0 | 194.9 | 144.2 | 75.3 | 146.7 | 90.3 | 128.6 | 108.1 |
| $\mathbf{5}$ | 177.5 | 183.2 | 210.5 | 168.1 | 84.1 | 130.1 | 43.1 | 45.3 | 55.9 | 71.7 | 68.4 |
| $\mathbf{6}$ | 46.5 | 139.8 | 102.8 | 118.4 | 60.1 | 65.3 | 39.9 | 23.8 | 15.5 | 38.3 | 39.1 |
| $\mathbf{7}$ | 13.2 | 52.7 | 63.9 | 49.5 | 32.9 | 30.7 | 21.2 | 15.4 | 9.5 | 13.8 | 18.3 |
| $\mathbf{8 +}$ | 4.9 | 22.6 | 24.5 | 33.1 | 20.5 | 25.1 | 24.1 | 14.1 | 6.1 | 10.7 | 6.7 |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 150.3 | 53.5 | 243.6 | 106.9 | 4.3 |
| $\mathbf{1}$ | 659.1 | 126.9 | 457.8 | 305.2 | 134.4 |
| $\mathbf{2}$ | 281.8 | 264.9 | 197.8 | 319.2 | 184.1 |
| $\mathbf{3}$ | 321.3 | 161.3 | 164.8 | 177.8 | 228.5 |
| $\mathbf{4}$ | 172.3 | 189.4 | 93.2 | 130.4 | 149.0 |
| $\mathbf{5}$ | 57.2 | 103.6 | 91.2 | 60.6 | 97.4 |
| $\mathbf{6}$ | 38.5 | 29.1 | 49.0 | 65.7 | 42.0 |
| $\mathbf{7}$ | 13.8 | 17.5 | 14.9 | 31.2 | 32.2 |
| $\mathbf{8 +}$ | 8.3 | 8.8 | 11.0 | 12.6 | 17.3 |

Table 3.6.2 WESTERN BALTIC HERRING. Input to ICA. Mean weight in catch (kg)

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 0.02957 | 0.01519 | 0.01535 | 0.01458 | 0.01010 | 0.01056 | 0.02962 | 0.01426 | 0.01112 | 0.02113 |
| $\mathbf{1}$ | 0.03476 | 0.03447 | 0.02545 | 0.03704 | 0.02092 | 0.02458 | 0.02748 | 0.03333 | 0.03433 | 0.02550 |
| $\mathbf{2}$ | 0.06685 | 0.06732 | 0.06797 | 0.08328 | 0.06843 | 0.08090 | 0.06845 | 0.06634 | 0.06583 | 0.05775 |
| $\mathbf{3}$ | 0.09490 | 0.09435 | 0.10204 | 0.10323 | 0.09841 | 0.09702 | 0.11807 | 0.09423 | 0.09814 | 0.09501 |
| $\mathbf{4}$ | 0.12342 | 0.11630 | 0.11428 | 0.12213 | 0.12349 | 0.11254 | 0.13420 | 0.11779 | 0.11642 | 0.13013 |
| $\mathbf{5}$ | 0.13901 | 0.14169 | 0.13615 | 0.14115 | 0.15196 | 0.13283 | 0.16198 | 0.13673 | 0.14713 | 0.14280 |
| $\mathbf{6}$ | 0.15560 | 0.16511 | 0.16795 | 0.15648 | 0.17041 | 0.13687 | 0.18170 | 0.16628 | 0.15660 | 0.14633 |
| $\mathbf{7}$ | 0.17091 | 0.17576 | 0.18228 | 0.17046 | 0.20626 | 0.15425 | 0.19671 | 0.16523 | 0.15382 | 0.15829 |
| $\mathbf{8 +}$ | 0.18256 | 0.19152 | 0.19890 | 0.18596 | 0.21696 | 0.19100 | 0.20872 | 0.18701 | 0.15756 | 0.15908 |
| 0.15560 |  |  |  |  |  |  |  |  |  |  |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 0.01053 | 0.01325 | 0.00618 | 0.01401 | 0.02116 |
| $\mathbf{1}$ | 0.02127 | 0.03152 | 0.02754 | 0.02719 | 0.03744 |
| $\mathbf{2}$ | 0.06998 | 0.06711 | 0.06419 | 0.07208 | 0.07306 |
| $\mathbf{3}$ | 0.09678 | 0.09075 | 0.10017 | 0.09378 | 0.09820 |
| $\mathbf{4}$ | 0.11956 | 0.10792 | 0.10596 | 0.11057 | 0.11521 |
| $\mathbf{5}$ | 0.14003 | 0.12234 | 0.13139 | 0.12280 | 0.15337 |
| $\mathbf{6}$ | 0.18763 | 0.13188 | 0.15228 | 0.14933 | 0.15760 |
| $\mathbf{7}$ | 0.18141 | 0.16029 | 0.16768 | 0.16192 | 0.18658 |
| $\mathbf{8 +}$ | 0.17170 | 0.16252 | 0.15295 | 0.17355 | 0.18501 |

Table 3.6.3 WESTERN BALTIC HERRING. Input to ICA.
Mean weight in stock (kg)

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 0.00010 | 0.00010 | 0.00010 | 0.00010 | 0.00010 | 0.00010 | 0.00010 | 0.00010 | 0.00010 | 0.00010 | 0.00010 |
| $\mathbf{1}$ | 0.03085 | 0.02029 | 0.01563 | 0.01855 | 0.01305 | 0.01815 | 0.01310 | 0.02209 | 0.02106 | 0.01398 | 0.01686 |
| $\mathbf{2}$ | 0.05277 | 0.04513 | 0.04020 | 0.05288 | 0.04590 | 0.05456 | 0.05147 | 0.05578 | 0.05668 | 0.04313 | 0.05088 |
| $\mathbf{3}$ | 0.07873 | 0.08176 | 0.09671 | 0.08357 | 0.07081 | 0.09051 | 0.10633 | 0.08293 | 0.08705 | 0.08370 | 0.07829 |
| $\mathbf{4}$ | 0.10412 | 0.10751 | 0.10793 | 0.10767 | 0.13269 | 0.11703 | 0.13334 | 0.11280 | 0.10813 | 0.12504 | 0.11594 |
| $\mathbf{5}$ | 0.12447 | 0.13127 | 0.14087 | 0.13921 | 0.16745 | 0.11974 | 0.16618 | 0.13378 | 0.14801 | 0.14365 | 0.16904 |
| $\mathbf{6}$ | 0.14492 | 0.15934 | 0.16715 | 0.15656 | 0.18923 | 0.15383 | 0.19429 | 0.16779 | 0.16015 | 0.16287 | 0.17627 |
| $\mathbf{7}$ | 0.15943 | 0.17102 | 0.18273 | 0.17676 | 0.20970 | 0.14667 | 0.20895 | 0.16832 | 0.14394 | 0.16503 | 0.16808 |
| $\mathbf{8 +}$ | 0.16398 | 0.18693 | 0.18906 | 0.20275 | 0.23377 | 0.12803 | 0.22635 | 0.18432 | 0.15043 | 0.18311 | 0.18052 |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 0.00010 | 0.00010 | 0.00010 | 0.00010 | 0.00010 |
| $\mathbf{1}$ | 0.01645 | 0.01444 | 0.01306 | 0.01260 | 0.01846 |
| $\mathbf{2}$ | 0.06368 | 0.04447 | 0.04561 | 0.05136 | 0.06210 |
| $\mathbf{3}$ | 0.09046 | 0.07926 | 0.08106 | 0.08000 | 0.09527 |
| $\mathbf{4}$ | 0.12388 | 0.10509 | 0.10925 | 0.10657 | 0.11740 |
| $\mathbf{5}$ | 0.17365 | 0.12681 | 0.14399 | 0.13221 | 0.16593 |
| $\mathbf{6}$ | 0.19830 | 0.15061 | 0.16285 | 0.15733 | 0.17102 |
| $\mathbf{7}$ | 0.19801 | 0.17287 | 0.19321 | 0.16766 | 0.18584 |
| $\mathbf{8 +}$ | 0.20363 | 0.18471 | 0.20759 | 0.18205 | 0.18708 |

Table 3.6.4 WESTERN BALTIC HERRING. Input to ICA.
Natural mortality

| Years | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age 8+ |  |  |  |  |  |  |  |  |
| 1991-2006 | 0.3 | 0.5 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |

Table 3.6.5 a
WESTERN BALTIC HERRING. Input to ICA.
AGE - STRUCTURED INDICES.
Fleet 1b: Acoustic Survey in Div. Illa+IVaE, Ages 2-8+
Ages 2-8+ (Catch: Number in millions)

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{2}$ | 1864.0 | 2092.0 | 2768.0 | 413.0 | 1887.0 | 1005.0 | 715.0 | 1682.0 | - | 1891.1 | 641.2 |
| $\mathbf{3}$ | 1927.0 | 1799.0 | 1274.0 | 935.0 | 1022.0 | 247.0 | 787.0 | 901.0 | - | 673.6 | 452.3 |
| $\mathbf{4}$ | 866.0 | 1593.0 | 598.0 | 501.0 | 1270.0 | 141.0 | 166.0 | 282.0 | - | 363.9 | 153.1 |
| $\mathbf{5}$ | 350.0 | 556.0 | 434.0 | 239.0 | 255.0 | 119.0 | 67.0 | 111.0 | - | 185.7 | 96.4 |
| $\mathbf{6}$ | 88.0 | 197.0 | 154.0 | 186.0 | 174.0 | 37.0 | 69.0 | 51.0 | - | 55.6 | 37.6 |
| $\mathbf{7}$ | 72.0 | 122.0 | 63.0 | 62.0 | 39.0 | 20.0 | 80.0 | 31.0 | - | 6.9 | 23.0 |
| $\mathbf{8 +}$ | 10.0 | 20.0 | 13.0 | 34.0 | 21.0 | 13.0 | 77.0 | 53.0 | - | 9.6 | 11.9 |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{2}$ | 1576.6 | 1110.0 | 929.6 | 1342.1 | 2217.0 |
| $\mathbf{3}$ | 1392.8 | 394.6 | 726.0 | 463.5 | 1780.4 |
| $\mathbf{4}$ | 524.3 | 323.4 | 306.9 | 201.3 | 490.0 |
| $\mathbf{5}$ | 87.5 | 103.4 | 183.7 | 102.5 | 180.4 |
| $\mathbf{6}$ | 39.5 | 25.2 | 72.1 | 83.6 | 27.0 |
| $\mathbf{7}$ | 17.8 | 12.0 | 21.5 | 37.2 | 9.5 |
| $\mathbf{8 +}$ | 17.1 | 5.4 | 18.0 | 21.4 | 0.1 |

Table 3.6.5 b WESTERN BALTIC HERRING. Input to ICA.
AGE - STRUCTURED INDICES.
Fleet 2b: Acoustic Survey in SD 22-24
Ages 0-5 (Catch: Number in millions)

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 5577.0 | 3467.0 | 768.0 | 4383.0 | 4001.0 | 1418.0 | 2608.0 | 2179.0 | 4821.0 | 1021.0 | 1831.0 |
| $\mathbf{1}$ | 2507.0 | 2179.0 | 345.0 | 412.0 | 1163.0 | 1084.0 | 1389.0 | 451.0 | 1145.0 | 1208.0 | 1314.0 |
| $\mathbf{2}$ | 880.0 | 1015.0 | 354.0 | 823.0 | 307.0 | 541.0 | 492.0 | 557.0 | 246.0 | 477.0 | 1761.0 |
| $\mathbf{3}$ | 852.0 | 465.0 | 485.0 | 540.0 | 332.0 | 413.0 | 343.0 | 364.0 | 187.0 | 348.0 | 1013.0 |
| $\mathbf{4}$ | 259.0 | 233.0 | 381.0 | 433.0 | 342.0 | 282.0 | 151.0 | 232.0 | 129.0 | 206.0 | 357.0 |
| $\mathbf{5}$ | 102.0 | 71.0 | 121.0 | 182.0 | 247.0 | 283.0 | 112.0 | 99.0 | 44.0 | 81.0 | 92.0 |
| $\mathbf{A g e}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |  |  |  |  |  |  |
| $\mathbf{0}$ | 3984.0 | 3701.0 | 2401.0 | 2769.0 | 3437.7 |  |  |  |  |  |  |
| $\mathbf{1}$ | 611.0 | 781.0 | 912.0 | 662.0 | 716.3 |  |  |  |  |  |  |
| $\mathbf{2}$ | 372.0 | 200.0 | 590.0 | 569.0 | 201.1 |  |  |  |  |  |  |
| $\mathbf{3}$ | 566.0 | 230.0 | 352.0 | 378.0 | 327.5 |  |  |  |  |  |  |
| $\mathbf{4}$ | 337.0 | 276.0 | 166.0 | 183.0 | 339.9 |  |  |  |  |  |  |
| $\mathbf{5}$ | 61.0 | 103.0 | 145.0 | 102.0 | 179.8 |  |  |  |  |  |  |

Table 3.6.5 c WESTERN BALTIC HERRING. Input to ICA.
AGE - STRUCTURED INDICES.
Fleet 4: IBTS in Kattegat
Ages 1-5 (Catch: Number per hour)

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 141.2 | 371.5 | 404.0 | 264.5 | 687.3 | 631.3 | 52.4 | 117.5 | 292.0 | - | 313.0 |
| $\mathbf{2}$ | 83.2 | 107.6 | 158.7 | 229.4 | 191.5 | 321.8 | 122.2 | 85.8 | 116.3 | - | 190.0 |
| $\mathbf{3}$ | 100.9 | 69.9 | 41.9 | 154.2 | 113.2 | 30.8 | 33.2 | 22.4 | 71.2 | - | 72.0 |
| $\mathbf{4}$ | 41.2 | 63.0 | 36.0 | 49.0 | 99.1 | 17.5 | 8.4 | 27.3 | 33.6 | - | 18.0 |
| $\mathbf{5}$ | 23.8 | 24.7 | 25.1 | 35.7 | 29.4 | 11.3 | 13.2 | 5.0 | 14.3 | - | 2.0 |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 1567.8 | 968.8 | 1225.2 | 607.2 | 508.6 |
| $\mathbf{2}$ | 169.0 | 550.2 | 215.0 | 255.4 | 78.8 |
| $\mathbf{3}$ | 100.2 | 170.2 | 143.6 | 53.7 | 63.6 |
| $\mathbf{4}$ | 15.5 | 52.7 | 30.0 | 23.3 | 40.1 |
| $\mathbf{5}$ | 5.8 | 29.4 | 23.0 | 12.5 | 31.9 |

# Table 3.6.6 WESTERN BALTIC HERRING: Input parameters for ICA FINAL Run 2007 

Integrated Catch at Age Analysis
--------------------------
Version 1.4 w
K.R.Patterson

Fisheries Research Services Marine Laboratory Aberdeen

24 August 1999
Type * to change language
Enter the name of the index file -->index.dat canum.low
weca.low
Stock weights in 2007 used for the year 2006 west.low
Natural mortality in 2007 used for the year 2006
natmor.low
Maturity ogive in 2007 used for the year 2006
matprop.low
Name of age-structured index file (Enter if none) : -->dagaiyfd.dat
Name of the SSB index file (Enter if none) -->
No indices of spawning biomass to be used.
No of years for separable constraint ?--> 5
Reference age for separable constraint ?--> 4
Constant selection pattern model (Y/N) ?-->y
$S$ to be fixed on last age ?--> 1.000000000000000
First age for calculation of reference F ?--> 3
Last age for calculation of reference $F$ ?--> 6
Use default weighting (Y/N) ?-->n
Enter relative weights at age
Weight for age $0-->\quad 0.100000000000000$
Weight for age 1--> 1.000000000000000
Weight for age 2--> 1.000000000000000
Weight for age 3--> 1.000000000000000
Weight for age 4--> 1.000000000000000
Weight for age 5--> 1.000000000000000
Weight for age 6--> 1.000000000000000
Weight for age 7--> 1.000000000000000
Weight for age 8--> 1.000000000000000
Enter relative weights by year
Weight for year 2002--> 1.000000000000000
Weight for year 2003--> 1.000000000000000
Weight for year 2004--> 1.000000000000000
Weight for year 2005--> 1.000000000000000
Weight for year 2006--> 1.000000000000000
Enter new weights for specified years and ages if needed
Enter year, age, new weight or -1,-1,-1 to end. -1 -1 -1.000000000000000
Is the last age of Acoustic Survey in Div IIIa+IVaE Ages a plus-group (Y/--
$>y$
Is the last age of Acoustic Survey in Sub div 22-24 Ages 0-a plus-group (Y--
$>n$
Is the last age of IYFS Katt Quart3 Age groups 1-5 (Mean Ca a plus-group (Y-$>n$
You must choose a catchability model for each index.
Models: A Absolute: Index = Abundance . e
L Linear: Index $=$ Q. Abundance . e
$P$ Power: Index $=$ Q. Abundance^ $K$.e
where Q and K are parameters to be estimated, and
e is a lognormally-distributed error.
Model for Acoustic Survey in Div IIIa+IVaE Ages is to be A/L/P ?-->L
Model for Acoustic Survey in Sub div 22-24 Ages 0- is to be A/L/P ?-->L
Model for IYFS Katt Quart3 Age groups 1-5 (Mean Ca is to be A/L/P ?-->L
Fit a stock-recruit relationship (Y/N) ?-->n
Enter lowest feasible F--> 5.0000000000000003E-02
Enter highest feasible F--> 1.000000000000000

Table 3.6.6 continued


No of years for separable analysis : 5
Age range in the analysis : 0 . . . 8
Year range in the analysis : 1991 . . . 2006
Number of indices of SSB : 0
Number of age-structured indices : 3
Parameters to estimate : 41
Number of observations : 316
Conventional single selection vector model to be fitted.


Table 3.6.7 WESTERN BALTIC HERRING. Output from ICA Final Run. FISHING MORTALITY (per year)

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 0.02820 | 0.04760 | 0.08110 | 0.05110 | 0.16960 | 0.04670 | 0.11940 | 0.12120 | 0.11100 | 0.05440 |
| $\mathbf{1}$ | 0.26280 | 0.17630 | 0.30290 | 0.16340 | 0.65010 | 0.38500 | 0.31930 | 0.36680 | 0.23880 | 0.33180 |
| $\mathbf{2}$ | 0.32970 |  |  |  |  |  |  |  |  |  |
| $\mathbf{2}$ | 0.32570 | 0.37880 | 0.35690 | 0.43090 | 0.58900 | 0.38950 | 0.36730 | 0.42940 | 0.42020 | 0.37430 |
| $\mathbf{3}$ | 0.43040 | 0.38080 | 0.48900 | 0.51600 | 0.52470 | 0.59160 | 0.51300 | 0.41900 | 0.46850 | 0.44000 |
| $\mathbf{4}$ | 0.41900 | 0.49170 | 0.51730 | 0.68110 | 0.54460 | 0.76250 | 0.46570 | 0.56790 | 0.37690 | 0.48010 |
| $\mathbf{5}$ | 0.40960 | 0.53690 | 0.66840 | 0.75740 | 0.47290 | 0.88420 | 0.54360 | 0.57150 | 0.44120 | 0.58480 |
| $\mathbf{6}$ | 0.29090 | 0.66370 | 0.66540 | 1.04840 | 0.68340 | 0.84310 | 0.76280 | 0.66370 | 0.38950 | 0.62000 |
| $\mathbf{7}$ | 0.56950 | 0.62460 | 0.74500 | 0.80950 | 0.99150 | 0.94030 | 0.74690 | 0.77450 | 0.61500 | 0.72330 |
| $\mathbf{8 +}$ | 0.56950 | 0.62460 | 0.74500 | 0.80950 | 0.99150 | 0.94030 | 0.74690 | 0.77450 | 0.61500 | 0.72330 |
| 0.69590 |  |  |  |  |  |  |  |  |  |  |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 0.03390 | 0.02660 | 0.02640 | 0.02930 | 0.03100 |
| $\mathbf{1}$ | 0.20230 | 0.15900 | 0.15750 | 0.17530 | 0.18520 |
| $\mathbf{2}$ | 0.31320 | 0.24620 | 0.24390 | 0.27140 | 0.28680 |
| $\mathbf{3}$ | 0.39300 | 0.30890 | 0.30610 | 0.34050 | 0.35990 |
| $\mathbf{4}$ | 0.65220 | 0.51270 | 0.50800 | 0.56520 | 0.59720 |
| $\mathbf{5}$ | 0.55940 | 0.43980 | 0.43570 | 0.48480 | 0.51230 |
| $\mathbf{6}$ | 0.67570 | 0.53120 | 0.52630 | 0.58550 | 0.61870 |
| $\mathbf{7}$ | 0.65220 | 0.51270 | 0.50800 | 0.56520 | 0.59720 |
| $\mathbf{8 +}$ | 0.65220 | 0.51270 | 0.50800 | 0.56520 | 0.59720 |

Table 3.6.8
WESTERN BALTIC HERRING. Output from ICA Final Run.
POPULATION ABUNDANCE ( millions)- 1 January

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 4948.6 | 3608.4 | 3056.6 | 6107.6 | 3999.8 | 4339.1 | 3863.8 | 5560.2 | 6257.6 | 3328.2 | 4613.7 |
| $\mathbf{1}$ | 4493.3 | 3564.1 | 2548.9 | 2088.0 | 4299.2 | 2500.9 | 3067.9 | 2540.2 | 3649.1 | 4148.7 | 2334.9 |
| $\mathbf{2}$ | 2137.0 | 2095.6 | 1812.3 | 1142.0 | 1075.5 | 1361.2 | 1032.2 | 1352.2 | 1067.7 | 1743.1 | 1805.8 |
| $\mathbf{3}$ | 1767.8 | 1263.3 | 1174.7 | 1038.4 | 607.7 | 488.6 | 754.9 | 585.3 | 720.6 | 574.2 | 981.5 |
| $\mathbf{4}$ | 895.5 | 941.1 | 706.8 | 589.8 | 507.5 | 294.4 | 221.4 | 370.0 | 315.2 | 369.3 | 302.8 |
| $\mathbf{5}$ | 578.8 | 482.2 | 471.2 | 345.0 | 244.4 | 241.0 | 112.4 | 113.8 | 171.7 | 177.0 | 187.1 |
| $\mathbf{6}$ | 202.2 | 314.6 | 230.8 | 197.8 | 132.4 | 124.7 | 81.5 | 53.5 | 52.6 | 90.4 | 80.8 |
| $\mathbf{7}$ | 33.3 | 123.8 | 132.6 | 97.1 | 56.7 | 54.7 | 43.9 | 31.1 | 22.5 | 29.2 | 39.8 |
| $\mathbf{8 +}$ | 12.4 | 53.1 | 50.9 | 64.9 | 35.3 | 44.8 | 50.0 | 28.5 | 14.5 | 22.6 | 14.5 |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 2812.7 | 4325.3 | 2900.0 | 2054.5 | 1313.5 | 1408.2 |
| $\mathbf{1}$ | 2772.8 | 2014.3 | 3120.1 | 2092.5 | 1478.0 | 943.3 |
| $\mathbf{2}$ | 1018.4 | 1373.8 | 1042.2 | 1616.6 | 1065.1 | 744.9 |
| $\mathbf{3}$ | 1039.5 | 609.6 | 879.3 | 668.6 | 1009.0 | 654.6 |
| $\mathbf{4}$ | 572.0 | 574.5 | 366.4 | 530.1 | 389.4 | 576.4 |
| $\mathbf{5}$ | 151.0 | 243.9 | 281.7 | 180.5 | 246.6 | 175.5 |
| $\mathbf{6}$ | 91.9 | 70.7 | 128.6 | 149.2 | 91.0 | 121.0 |
| $\mathbf{7}$ | 31.2 | 38.3 | 34.0 | 62.2 | 68.0 | 40.1 |
| $\mathbf{8 +}$ | 19.0 | 24.1 | 30.3 | 32.0 | 42.0 | 49.6 |

Table 3.6.9 WESTERN BALTIC HERRING. Output from ICA Final Run. STOCK SUMMARY

| Year | Recruits <br> Age 0 <br> (thousands) | Total <br> Biomass <br> (tonnes) | Spawning <br> Biomass <br> (tonnes) | Landings | Yield <br> SBB <br> ratio | Mean F <br> Ages <br> 3-6 | SoP |
| :--- | :--- | :--- | :--- | :--- | ---: | ---: | ---: |
| $\mathbf{1 9 9 1}$ | 4948600 | 592997 | 291885 | 191573 | 0.6563 | 0.3875 | 99 |
| $\mathbf{1 9 9 2}$ | 3608410 | 516235 | 300193 | 194411 | 0.6476 | 0.5183 | 100 |
| $\mathbf{1 9 9 3}$ | 3056560 | 441711 | 275864 | 185010 | 0.6707 | 0.5850 | 100 |
| $\mathbf{1 9 9 4}$ | 6107580 | 359334 | 216065 | 172438 | 0.7981 | 0.7507 | 99 |
| $\mathbf{1 9 9 5}$ | 3999770 | 302368 | 169325 | 150831 | 0.8908 | 0.5564 | 100 |
| $\mathbf{1 9 9 6}$ | 4339120 | 260568 | 124469 | 121266 | 0.9743 | 0.7704 | 100 |
| $\mathbf{1 9 9 7}$ | 3863760 | 258511 | 137152 | 115588 | 0.8428 | 0.5713 | 100 |
| $\mathbf{1 9 9 8}$ | 5560200 | 257062 | 111615 | 107032 | 0.9589 | 0.5555 | 99 |
| $\mathbf{1 9 9 9}$ | 6257570 | 274058 | 117525 | 97240 | 0.8274 | 0.4190 | 100 |
| $\mathbf{2 0 0 0}$ | 3328150 | 276864 | 128178 | 109914 | 0.8575 | 0.5312 | 100 |
| $\mathbf{2 0 0 1}$ | 4613680 | 298829 | 147878 | 105803 | 0.7155 | 0.5239 | 99 |
| $\mathbf{2 0 0 2}$ | 2812740 | 330136 | 182038 | 106191 | 0.5833 | 0.5701 | 99 |
| $\mathbf{2 0 0 3}$ | 4325300 | 251938 | 141617 | 78309 | 0.5530 | 0.4482 | 99 |
| $\mathbf{2 0 0 4}$ | 2900010 | 274243 | 158180 | 76815 | 0.4856 | 0.4440 | 100 |
| $\mathbf{2 0 0 5}$ | 2054450 | 283160 | 155248 | 88406 | 0.5694 | 0.4940 | 100 |
| $\mathbf{2 0 0 6}$ | 1313450 | 312382 | 184516 | 88931 | 0.4820 | 0.5220 | 100 |

Table 3.6.10

## WESTERN BALTIC HERRING. Output from ICA Final Run.

 PARAMETER ESTIMATES| Parm. No. |  | Max. <br> Likelh. <br> Estim. | Cv | Lower 95\% CL | $\begin{gathered} \text { Upper } \\ 95 \% \mathrm{CL} \end{gathered}$ | -s.e. | +s.e. | Mean of param. Estim. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Separable model : F by year |  |  |  |  |  |  |  |  |
| 1 | 2002 | 0.6522 | 14 | 0.4929 | 0.8632 | 0.5654 | 0.7525 | 0.6589 |
| 2 | 2003 | 0.5127 | 14 | 0.3852 | 0.6825 | 0.4431 | 0.5933 | 0.5182 |
| 3 | 2004 | 0.5080 | 14 | 0.3807 | 0.6778 | 0.4385 | 0.5885 | 0.5135 |
| 4 | 2005 | 0.5652 | 15 | 0.4166 | 0.7667 | 0.4837 | 0.6603 | 0.572 |
| 5 | 2006 | 0.5972 | 18 | 0.4162 | 0.857 | 0.4967 | 0.7181 | 0.6074 |
| Separable Model: Selection (S) by age |  |  |  |  |  |  |  |  |
| 6 | 0 | 0.0519 | 42 | 0.0224 | 0.1203 | 0.0338 | 0.0797 | 0.0569 |
| 7 | 1 | 0.3101 | 18 | 0.2175 | 0.4422 | 0.2588 | 0.3716 | 0.3152 |
| 8 | 2 | 0.4802 | 17 | 0.3417 | 0.6747 | 0.4037 | 0.5712 | 0.4875 |
| 9 | 3 | 0.6026 | 17 | 0.4309 | 0.8425 | 0.5078 | 0.715 | 0.6114 |
| 4 | 1 | Fixed: | nce |  |  |  |  |  |
| 10 | 5 | 0.8577 | 15 | 0.6361 | 1.1566 | 0.7364 | 0.9991 | 0.8678 |
| 11 | 6 | 1.0360 | 14 | 0.7765 | 1.3822 | 0.8943 | 1.2002 | 1.0473 |
| 7 | 1 | Fixed: | ag |  |  |  |  |  |
| Separable Model: Populations in year 2007 |  |  |  |  |  |  |  |  |
| 12 | 0 | 1313454 | 57 | 424063 | 4068172 | 737756 | 2338391 | 2E+06 |
| 13 | 1 | 1477962 | 25 | 894493 | 2442023 | 1143918 | 1909554 | 2E+06 |
| 14 | 2 | 1065115 | 18 | 735017 | 1543460 | 881461 | 1287033 | 1E+06 |
| 15 | 3 | 1008998 | 16 | 734936 | 1385259 | 858352 | 1186083 | 1E+06 |
| 16 | 4 | 389393 | 14 | 292578 | 518244 | 336549 | 450534 | 393556 |
| 17 | 5 | 246630 | 14 | 184403 | 329857 | 212627 | 286071 | 249359 |
| 18 | 6 | 91023 | 16 | 66220 | 125115 | 77386 | 107063 | 92230 |
| 19 | 7 | 67998 | 19 | 46641 | 99133 | 56100 | 82419 | 69267 |
| Separable Model: Populations at age |  |  |  |  |  |  |  |  |
| 20 | 2002 | 31242 | 27 | 18191 | 53654 | 23708 | 41168 | 32454 |
| 21 | 2003 | 38287 | 22 | 24705 | 59334 | 30618 | 47876 | 39255 |
| 22 | 2004 | 34019 | 19 | 23352 | 49559 | 28078 | 41219 | 34652 |
| 23 | 2005 | 62225 | 18 | 43132 | 89770 | 51613 | 75019 | 63322 |

Table 3.6.11

## WESTERN BALTIC HERRING. Output from ICA Final Run. AGE STRUCTURED INDEX OF CATCHABILITIES

Fleet 1b: Acoustic Survey in Div. Illa+IVaE, Ages 2-8+
Linear model fitted. Slopes at age :

| 24 | 2 Q | 1332.0 | 19 | 1108.0 | 2355.0 | 1332.0 | 1957.0 | 1645.0 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 25 | 3 Q | 1474.0 | 19 | 1226.0 | 2604.0 | 1474.0 | 2165.0 | 1820.0 |
| 26 | $4 Q$ | 1354.0 | 19 | 1126.0 | 2393.0 | 1354.0 | 1989.0 | 1672.0 |
| 27 | 5 Q | 1125.0 | 19 | 934.5 | 1992.0 | 1125.0 | 1655.0 | 1390.0 |
| 28 | 6 Q | 977.4 | 19 | 810.4 | 1741.0 | 977.4 | 1444.0 | 1211.0 |
| 29 | 7 Q | 1036.0 | 19 | 856.1 | 1863.0 | 1036.0 | 1540.0 | 1288.0 |
| 30 | 8 Q | 710.8 | 19 | 589.3 | 1267.0 | 710.8 | 1051.0 | 880.9 |

## Fleet 2b: Acoustic Survey in SD 22-24

Linear model fitted. Slopes at age :

| 31 | 0 Q | 975.6 | 17 | 823.0 | 1648.0 | 975.6 | 1391.0 | 1183.0 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 32 | 1 Q | 607.6 | 17 | 514.4 | 1016.0 | 607.6 | 859.7 | 733.7 |
| 33 | 2 Q | 562.7 | 17 | 476.7 | 938.5 | 562.7 | 795.0 | 678.9 |
| 34 | 3 Q | 813.7 | 17 | 689.5 | 1356.0 | 813.7 | 1149.0 | 981.5 |
| 35 | 4 Q | 1003.0 | 17 | 849.8 | 1674.0 | 1003.0 | 1418.0 | 1211.0 |
| 36 | 5 Q | 872.3 | 17 | 738.3 | 1459.0 | 872.3 | 1235.0 | 1054.0 |

## Fleet 4: IBTS in Kattegat

Linear model fitted. Slopes at age

| 37 | 1 | $Q$ | 0.00023 | 16 | 0.00020 | 0.00038 | 0.00023 | 0.00032 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.00028 |  |  |  |  |  |  |  |  |
| 38 | $2 ~ Q ~$ | 0.00017 | 16 | 0.00015 | 0.00028 | 0.00017 | 0.00024 | 0.00021 |
| 39 | $3 Q$ | 0.00012 | 16 | 0.00010 | 0.00020 | 0.00012 | 0.00017 | 0.00015 |
| 40 | $4 Q$ | 0.00011 | 16 | 0.00009 | 0.00017 | 0.00011 | 0.00015 | 0.00013 |
| 41 | $5 Q$ | 0.00010 | 16 | 0.00009 | 0.00016 | 0.00010 | 0.00014 | 0.00012 |

Table 3.6.12
WESTERN BALTIC HERRING. Output from ICA Final Run. RESIDUALS ABOUT THE MODEL FIT Separable Model Residuals (log(Observed Catch) $\log ($ Expected Catch))

| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 0.618 | -0.608 | 1.317 | 0.733 | -2.086 |
| $\mathbf{1}$ | 0.492 | -0.615 | 0.240 | 0.135 | -0.388 |
| $\mathbf{2}$ | 0.122 | -0.030 | -0.037 | -0.091 | -0.273 |
| $\mathbf{3}$ | 0.042 | 0.088 | -0.248 | 0.011 | -0.196 |
| $\mathbf{4}$ | -0.377 | -0.106 | -0.358 | -0.474 | -0.073 |
| $\mathbf{5}$ | -0.035 | 0.269 | 0.005 | -0.044 | 0.075 |
| $\mathbf{6}$ | -0.071 | 0.089 | 0.017 | 0.082 | 0.087 |
| $\mathbf{7}$ | 0.009 | 0.218 | 0.183 | 0.240 | 0.140 |

Table 3.6.13
WESTERN BALTIC HERRING. Output from ICA Final Run.
AGED INDEX RESIDUALS:
LOG(OBSERVED INDEX) LOG(EXPECTED INDEX

Fleet 1b: Acoustic Survey in Div. Illa+IVaE, Ages 2-8+

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{2}$ | -0.095 | 0.073 | 0.485 | -0.910 | 0.769 | -0.222 | -0.299 | 0.325 | $* * * * * * *$ | 0.154 | -0.977 |
| $\mathbf{3}$ | 0.092 | 0.328 | 0.123 | -0.046 | 0.585 | -0.576 | 0.099 | 0.430 | $* * * * * * *$ | 0.171 | -0.825 |
| $\mathbf{4}$ | 0.050 | 0.656 | -0.022 | 0.084 | 1.080 | -0.438 | -0.175 | -0.095 | $* * * * * * *$ | 0.107 | -0.550 |
| $\mathbf{5}$ | -0.240 | 0.485 | 0.343 | 0.114 | 0.345 | -0.146 | -0.171 | 0.340 | $* * * * * * *$ | 0.421 | -0.337 |
| $\mathbf{6}$ | -0.502 | 0.095 | 0.159 | 0.742 | 0.848 | -0.540 | 0.458 | 0.516 | $* * * * * * *$ | 0.049 | -0.148 |
| $\mathbf{7}$ | 1.216 | 0.466 | -0.189 | 0.147 | 0.335 | -0.329 | 1.156 | 0.570 | $* * * * * * *$ | -0.900 | -0.024 |
| $\mathbf{8 +}$ | 0.606 | -0.119 | -0.433 | 0.325 | 0.566 | -0.183 | 1.365 | 1.570 | $* * * * * * *$ | 0.061 | 0.700 |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{2}$ | 0.471 | -0.221 | -0.124 | -0.178 | 0.750 |
| $\mathbf{3}$ | 0.275 | -0.505 | -0.264 | -0.417 | 0.530 |
| $\mathbf{4}$ | 0.143 | -0.432 | -0.038 | -0.793 | 0.425 |
| $\mathbf{5}$ | -0.189 | -0.576 | -0.148 | -0.256 | 0.015 |
| $\mathbf{6}$ | -0.274 | -0.551 | -0.102 | -0.065 | -0.681 |
| $\mathbf{7}$ | -0.065 | -0.750 | -0.052 | -0.071 | -1.505 |
| $\mathbf{8 +}$ | 0.770 | -0.707 | 0.265 | 0.419 | -5.200 |

Fleet 2b: Acoustic Survey in SD 22-24

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 0.407 | 0.263 | -1.052 | -0.026 | 0.401 | -0.816 | -0.033 | -0.575 | 0.093 | -0.873 |
| $\mathbf{1}$ | 0.525 | 0.547 | -0.859 | -0.594 | 0.111 | 0.37 | 0.361 | -0.537 | -0.07 | -0.07 |
| $\mathbf{2}$ | 0.108 | 0.313 | -0.612 | 0.752 | -0.047 | 0.124 | 0.288 | 0.192 | -0.397 | -0.261 |
| $\mathbf{3}$ | -0.019 | -0.329 | -0.127 | 0.125 | 0.181 | 0.671 | -0.012 | 0.226 | -0.608 | 0.217 |
| $\mathbf{4}$ | -0.749 | -0.846 | -0.047 | 0.392 | 0.198 | 0.724 | 0.147 | 0.144 | -0.435 | -0.043 |
| $\mathbf{5}$ | -1.112 | -1.19 | -0.528 | 0.263 | 0.686 | 1.165 | 0.728 | 0.615 | -0.712 | -0.017 |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ | 0.64 | 0.13 | 0.097 | 0.587 | 1.252 |
| $\mathbf{1}$ | -0.452 | 0.078 | -0.206 | -0.112 | 0.322 |
| $\mathbf{2}$ | -0.022 | -0.995 | 0.361 | -0.092 | -0.703 |
| $\mathbf{3}$ | 0.073 | -0.361 | -0.304 | 0.068 | -0.470 |
| $\mathbf{4}$ | 0.15 | -0.166 | -0.229 | -0.455 | 0.499 |
| $\mathbf{5}$ | -0.163 | -0.214 | -0.019 | 0.114 | 0.391 |

Fleet 4: IBTS in Kattegat

| Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | -1.528 | -0.383 | 0.115 | -0.196 | 0.340 | 0.632 | -2.103 | -1.077 | -0.609 | $* * * * * *$ |
| $\mathbf{2}$ | -1.166 | -0.856 | -0.336 | 0.541 | 0.519 | 0.678 | -0.028 | -0.612 | -0.078 | $* * * * * * *$ |
| $\mathbf{3}$ | -0.362 | -0.424 | -0.795 | 0.648 | 0.880 | -0.163 | -0.571 | -0.771 | 0.210 | $* * * * * * *$ |
| $\mathbf{4}$ | -0.455 | -0.034 | -0.290 | 0.300 | 1.070 | 0.017 | -0.622 | 0.112 | 0.361 | $* * * * * * *$ |
| $\mathbf{5}$ | -0.514 | -0.217 | -0.094 | 0.623 | 0.596 | -0.090 | 0.616 | -0.356 | 0.210 | $* * * * * * *$ |


| Age | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 1.324 | 1.135 | 0.931 | 0.640 | 0.816 |
| $\mathbf{2}$ | 0.276 | 1.115 | 0.451 | 0.201 | -0.548 |
| $\mathbf{3}$ | 0.139 | 1.150 | 0.612 | -0.077 | -0.307 |
| $\mathbf{4}$ | -0.835 | 0.295 | 0.178 | -0.408 | 0.462 |
| $\mathbf{5}$ | -0.486 | 0.579 | 0.186 | 0.055 | 0.694 |

Table 3.6.14
WESTERN BALTIC HERRING. Output from ICA Final Run. PARAMETERS OF THE DISTRIBUTION OF In CATCHES AT AGE

| Separable model fitted from 2002 to 2006 |  |
| :--- | ---: |
| Variance | 0.1512 |
| Skewness test stat. | -2.1845 |
| Kurtosis test statistic | 0.5288 |
| Partial chi-square | 0.2212 |
| Significance in fit | 0.0000 |
| Degrees of freedom | 17 |

Table 3.6.15

## WESTERN BALTIC HERRING. Output from ICA Final Run. PARAMETERS OF THE DISTRIBUTION OF THE AGE-STRUCTURED INDICES

DISTRIBUTION STATISTICS FOR Acoustic Survey in Div IIla+IVaE WR 2-8+
Linear catchability relationship assumed

| Age | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Variance | 0.0385 | 0.0262 | 0.0323 | 0.0148 | 0.0327 | 0.0743 | 0.3493 |
| Skewness test stat. | -0.5093 | -0.7157 | 0.852 | 0.1159 | 0.4952 | -0.2902 | -4.0764 |
| Kurtosis test statisti | -0.3768 | -0.7109 | 0.1507 | -0.9021 | -0.7584 | -0.0801 | 5.1371 |
| Partial chi-square | 0.0258 | 0.0179 | 0.0228 | 0.0109 | 0.0255 | 0.0607 | 0.2942 |
| Significance in fit | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| Number of observations | 15 | 15 | 15 | 15 | 15 | 15 | 15 |
| Degrees of freedom | 14 | 14 | 14 | 14 | 14 | 14 | 14 |
| Weight in the analysis | 0.1429 | 0.1429 | 0.1429 | 0.1429 | 0.1429 | 0.1429 | 0.1429 |

DISTRIBUTION STATISTICS FOR Acoustic Survey in Subdiv 22-24 WR 0-5
Linear catchability relationship assumed

| Age | 0 | 1 | 2 | 3 | 4 | 5 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Variance | 0.0651 | 0.0329 | 0.0447 | 0.0222 | 0.0367 | 0.0720 |
| Skewness test stat. | -0.0479 | -0.5885 | -0.1057 | 0.4452 | -0.2911 | -0.3886 |
| Kurtosis test statisti | -0.4652 | -0.7655 | -0.2537 | -0.3694 | -0.6080 | -0.4639 |
| Partial chi-square | 0.0457 | 0.0239 | 0.0334 | 0.0169 | 0.0283 | 0.0576 |
| Significance in fit | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| Number of observations | 16 | 16 | 16 | 16 | 16 | 16 |
| Degrees of freedom | 15 | 15 | 15 | 15 | 15 | 15 |
| Weight in the analysis | 0.1667 | 0.1667 | 0.1667 | 0.1667 | 0.1667 | 0.1667 |

DISTRIBUTION STATISTICS FOR IYFS Kattegat Quarter 3 WR 1-5
Linear catchability relationship assumed

| Age | 1 | 2 | 3 | 4 | 5 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| Variance | 0.1998 | 0.0785 | 0.071 | 0.0468 | 0.085 |
| Skewness test stat. | -1.0212 | -0.2526 | 0.7759 | 0.4319 | -2.0693 |
| Kurtosis test statisti | -0.3838 | -0.5678 | -0.6522 | -0.0456 | 1.4464 |
| Partial chi-square | 0.4633 | 0.2096 | 0.2407 | 0.1913 | 0.4675 |
| Significance in fit | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| Number of observations | 15 | 15 | 15 | 15 | 15 |
| Degrees of freedom | 14 | 14 | 14 | 14 | 14 |
| Weight in the analysis | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |

Table 3.6.16
WESTERN BALTIC HERRING. Output from ICA Final Run. ANALYSIS OF VARIANCE TABLE

| Unweighted Statistics | SSQ | Data | Param. | d.f. | Variance |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Variance |  |  |  |  |  |
| Total for modeI | 123.1888 | 316 | 41 | 275 | 0.4480 |
| Catches at age | 9.2091 | 40 | 23 | 17 | 0.5417 |
| Aged Indices |  |  |  |  |  |
| Acoustic Survey Div IIIa+IVaE WR 2-8+ | 55.6831 | 105 | 7 | 98 | 0.5682 |
| Acoustic Survey Subdiv 22-24 WR 0-5 | 24.6146 | 96 | 6 | 90 | 0.2735 |
| IYFS Kattegat Quarter 3 WR 1-5 | 33.6821 | 75 | 5 | 70 | 0.4812 |


| Weighted Statistics | SSQ | Data | Param. | d.f. | Variance |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Variance |  |  |  |  |  |
| Total for model | 5.7385 | 316 | 41 | 275 | 0.0209 |
| Catches at age | 2.5711 | 40 | 23 | 17 | 0.1512 |
| Aged Indices |  |  |  |  |  |
| Acoustic Survey Div Illa+IVaE WR 2-8+ | 1.1364 | 105 | 7 | 98 | 0.0116 |
| Acoustic Survey Subdiv 22-24 WR 0-5 | 0.6837 | 96 | 6 | 90 | 0.0076 |
| IYFS Kattegat Quarter 3 WR 1-5 | 1.3473 | 75 | 5 | 70 | 0.0192 |

Table 3.7.1
WESTERN BALTIC HERRING. Input table for short term predictions

MFDP version 1a
Run: WBSS GeoMean 5 years
Time and date: 14:54 18/03/2007
Fbar age range: 3-6

| $\begin{aligned} & 2007 \\ & \text { Age } \end{aligned}$ | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 3521756 | 0.3 | 0.00 | 0.1 | 0.25 | 0.000 | 0.031 | 0.014 |
| 1 | 2432366 | 0.5 | 0.00 | 0.1 | 0.25 | 0.015 | 0.185 | 0.031 |
| 2 | 744900 | 0.2 | 0.20 | 0.1 | 0.25 | 0.053 | 0.287 | 0.070 |
| 3 | 654600 | 0.2 | 0.75 | 0.1 | 0.25 | 0.085 | 0.360 | 0.097 |
| 4 | 576400 | 0.2 | 0.90 | 0.1 | 0.25 | 0.111 | 0.597 | 0.111 |
| 5 | 175500 | 0.2 | 1.00 | 0.1 | 0.25 | 0.147 | 0.512 | 0.136 |
| 6 | 121000 | 0.2 | 1.00 | 0.1 | 0.25 | 0.164 | 0.619 | 0.153 |
| 7 | 40100 | 0.2 | 1.00 | 0.1 | 0.25 | 0.182 | 0.597 | 0.172 |
| 8 | 49600 | 0.2 | 1.00 | 0.1 | 0.25 | 0.192 | 0.597 | 0.171 |
| 2008 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 0 | 3521756 | 0.3 | 0.00 | 0.1 | 0.25 | 0.000 | 0.031 | 0.014 |
| 1 |  | 0.5 | 0.00 | 0.1 | 0.25 | 0.015 | 0.185 | 0.031 |
| 2 |  | 0.2 | 0.20 | 0.1 | 0.25 | 0.053 | 0.287 | 0.070 |
| 3 |  | 0.2 | 0.75 | 0.1 | 0.25 | 0.085 | 0.360 | 0.097 |
| 4 |  | 0.2 | 0.90 | 0.1 | 0.25 | 0.111 | 0.597 | 0.111 |
| 5 |  | 0.2 | 1.00 | 0.1 | 0.25 | 0.147 | 0.512 | 0.136 |
| 6 |  | 0.2 | 1.00 | 0.1 | 0.25 | 0.164 | 0.619 | 0.153 |
| 7 |  | 0.2 | 1.00 | 0.1 | 0.25 | 0.182 | 0.597 | 0.172 |
| 8 |  | 0.2 | 1.00 | 0.1 | 0.25 | 0.192 | 0.597 | 0.171 |
| 2009 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 0 | 3521756 | 0.3 | 0.00 | 0.1 | 0.25 | 0.000 | 0.031 | 0.014 |
| 1 |  | 0.5 | 0.00 | 0.1 | 0.25 | 0.015 | 0.185 | 0.031 |
| 2 |  | 0.2 | 0.20 | 0.1 | 0.25 | 0.053 | 0.287 | 0.070 |
| 3 |  | 0.2 | 0.75 | 0.1 | 0.25 | 0.085 | 0.360 | 0.097 |
| 4 |  | 0.2 | 0.90 | 0.1 | 0.25 | 0.111 | 0.597 | 0.111 |
| 5 |  | 0.2 | 1.00 | 0.1 | 0.25 | 0.147 | 0.512 | 0.136 |
| 6 |  | 0.2 | 1.00 | 0.1 | 0.25 | 0.164 | 0.619 | 0.153 |
| 7 |  | 0.2 | 1.00 | 0.1 | 0.25 | 0.182 | 0.597 | 0.172 |
| 8 |  | 0.2 | 1.00 | 0.1 | 0.25 | 0.192 | 0.597 | 0.171 |

Input units are thousands and kg - output in tonnes

| $\mathrm{M}=$ | Natural mortality |
| :--- | :--- |
| MAT $=$ | Maturity ogive |
| $\mathrm{PF}=$ | Proportion of F before spawning |
| $\mathrm{PM}=$ | Proportion of M before spawning |
| SWT $=$ | Weight in stock $(\mathrm{kg})$ |
| Sel $=$ | Exploit. Pattern |
| CWT $=$ | Weight in catch $(\mathrm{kg})$ |

$\mathrm{N}_{2007}$ Age 1: $\quad$ Geometric Mean from ICA of age 1 (Table 3.6.8) for the years 2001-2005
$\mathrm{N}_{2007}$ Age 2-8+
$\mathrm{N}_{\text {2006/2007/2008/2009 }}$ Age 0: Output from ICA (Table 3.6.8)
Geometric Mean from ICA of age 0 (Table 3.6.8) for the years 2000-2004
Natural Mortality (M): Average for 2004-2006

Weight in the Catch/Stock (CWt/SWt): Average for 2004-2006
Expoitation pattern (Sel):
Average for 2004-2006 rescaled to the last year

Table 3.7.2 WESTERN BALTIC HERRING.
Short term prediction single option table, status quo F.
MFDP version 1a
Run: WBSS GeoMean 5 years
Time and date: 14:54 18/03/2007
Fbar age range: 3-6

| Year: |  | 2007 F multiplie 1 |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age |  | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) SSNos(ST) | SSB(ST) |  |
|  | 0 | 0.031 | 92941 | 1281 | 3521756 | 352 | 0 | 0 | 0 | 0 |
|  | 1 | 0.1852 | 326103 | 10019 | 2432366 | 35772 | 0 | 0 | 0 | 0 |
|  | 2 | 0.2868 | 169132 | 11801 | 744900 | 39497 | 148980 | 7899 | 137708 | 7302 |
|  | 3 | 0.3599 | 180384 | 17566 | 654600 | 55931 | 490950 | 41948 | 450499 | 38492 |
|  | 4 | 0.5972 | 237242 | 26234 | 576400 | 64023 | 518760 | 57620 | 464852 | 51633 |
|  | 5 | 0.5123 | 64307 | 8736 | 175500 | 25865 | 175500 | 25865 | 158604 | 23375 |
|  | 6 | 0.6187 | 51116 | 7824 | 121000 | 19812 | 121000 | 19812 | 108193 | 17715 |
|  | 7 | 0.5972 | 16505 | 2840 | 40100 | 7308 | 40100 | 7308 | 35933 | 6548 |
|  | 8 | 0.5972 | 20415 | 3481 | 49600 | 9535 | 49600 | 9535 | 44446 | 8544 |
| Total |  |  | 1158146 | 89783 | 8316222 | 258094 | 1544890 | 169987 | 1400234 | 153608 |


| Year: <br> Age |  | 2008 F multiplic 1 |  | Fbar: 0.522 |  |  | SSNos(Jan) | SSB(Jan) SSNos(ST) |  | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | F | CatchNos | Yield | StockNos | Biomass |  |  |  |  |
|  | 0 | 0.031 | 92941 | 1281 | 3521756 | 352 | 0 | 0 | 0 | 0 |
|  | 1 | 0.1852 | 339105 | 10418 | 2529347 | 37198 | 0 | 0 | 0 | 0 |
|  | 2 | 0.2868 | 278340 | 19422 | 1225880 | 65000 | 245176 | 13000 | 226625 | 12016 |
|  | 3 | 0.3599 | 126157 | 12286 | 457816 | 39117 | 343362 | 29338 | 315071 | 26921 |
|  | 4 | 0.5972 | 153921 | 17021 | 373965 | 41537 | 336568 | 37384 | 301593 | 33499 |
|  | 5 | 0.5123 | 95164 | 12928 | 259711 | 38275 | 259711 | 38275 | 234708 | 34590 |
|  | 6 | 0.6187 | 36367 | 5567 | 86087 | 14095 | 86087 | 14095 | 76975 | 12603 |
|  | 7 | 0.5972 | 21963 | 3779 | 53360 | 9724 | 53360 | 9724 | 47815 | 8714 |
|  | 8 | 0.5972 | 16635 | 2836 | 40416 | 7770 | 40416 | 7770 | 36216 | 6962 |
| Total |  |  | 1160594 | 85538 | 8548338 | 253070 | 1364681 | 149586 | 1239004 | 135306 |


| Year: | 2009 F multiplie 1 |  |  |  |  |  |  |  |  | Fbar: |  |  |  |  |  |  |  | 0.522 |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |  |  |  |  |  |  |  |  |  |  |  |
|  | 0 | 0.031 | 92941 | 1281 | 3521756 | 352 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0.1852 | 339105 | 10418 | 2529347 | 37198 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2 | 0.2868 | 289438 | 20196 | 1274757 | 67592 | 254951 | 13518 | 235661 | 12496 |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.3599 | 207617 | 20218 | 753427 | 64375 | 565070 | 48281 | 518512 | 44303 |  |  |  |  |  |  |  |  |  |  |  |
|  | 4 | 0.5972 | 107650 | 11904 | 261545 | 29051 | 235390 | 26146 | 210929 | 23429 |  |  |  |  |  |  |  |  |  |  |  |
|  | 5 | 0.5123 | 61742 | 8388 | 168499 | 24833 | 168499 | 24833 | 152277 | 22442 |  |  |  |  |  |  |  |  |  |  |  |
|  | 6 | 0.6187 | 53818 | 8238 | 127394 | 20859 | 127394 | 20859 | 113911 | 18651 |  |  |  |  |  |  |  |  |  |  |  |
|  | 7 | 0.5972 | 15626 | 2689 | 37964 | 6918 | 37964 | 6918 | 34019 | 6199 |  |  |  |  |  |  |  |  |  |  |  |
|  | 8 | 0.5972 | 17391 | 2965 | 42253 | 8123 | 42253 | 8123 | 37863 | 7279 |  |  |  |  |  |  |  |  |  |  |  |
| Total |  |  | 1185327 | 86297 | 8716942 | 259301 | 1431522 | 148678 | 1303171 | 134799 |  |  |  |  |  |  |  |  |  |  |  |

[^7]Table 3.7.3 WESTERN BALTIC HERRING.
Short term prediction single option table
2007: Status quo $F$ and 2008/2009: $\mathrm{F}_{0.1}$
MFDP version 1a
Run: WBSS F0-1
Time and date: 16:36 19/03/2007
Fbar age range: 3-6


| Year: Age |  | 2009 F multiplie 0.4203 |  |  | Fbar: 0.2194 |  |  | SSB(Jan) | SSNos(ST) | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | F | CatchNos | Yield | StockNos | Biomass | Nos(Jan) |  |  |  |
|  | 0 | 0.0130 | 39398 | 543 | 3521756 | 352 | 0 | 0 | 0 | 0 |
|  | 1 | 0.0778 | 152259 | 4678 | 2575209 | 37873 | 0 | 0 | 0 | 0 |
|  | 2 | 0.1205 | 146362 | 10213 | 1419234 | 75253 | 283847 | 15051 | 266768 | 14145 |
|  | 3 | 0.1513 | 113476 | 11051 | 889694 | 76018 | 667270 | 57014 | 625199 | 53419 |
|  | 4 | 0.2510 | 65102 | 7199 | 322213 | 35789 | 289992 | 32210 | 269011 | 29880 |
|  | 5 | 0.2153 | 41972 | 5702 | 238207 | 35106 | 238207 | 35106 | 221762 | 32683 |
|  | 6 | 0.2601 | 35736 | 5470 | 171443 | 28071 | 171443 | 28071 | 158895 | 26016 |
|  | 7 | 0.2510 | 10980 | 1889 | 54342 | 9903 | 54342 | 9903 | 50411 | 9187 |
|  | 8 | 0.2510 | 12069 | 2058 | 59734 | 11483 | 59734 | 11483 | 55412 | 10652 |
| Total |  |  | 617355 | 48802 | 9251832 | 309849 | 1764834 | 188838 | 1647458 | 175982 |

Table 3.7.4 WESTERN BALTIC HERRING.
Short term prediction single option table
2007: Status quo F; 2008: Catch reduction of $15 \%$; 2009: Status quo $F$
MFDP version 1a
Run: WBSS -15\%
Time and date: 17:50 19/03/2007
Fbar age range: 3-6

| Year: Age | 2007 F multiplie 1 |  |  | Fbar: |  | 0.522 |  | SSB(Jan) | SSNos(ST) | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | F | CatchNos | Yield | StockNos | Biomass | Nos(Jan) |  |  |  |
|  | 0 | 0.0310 | 92941 | 1281 | 3521756 | 352 | 0 | 0 | 0 | 0 |
|  | 1 | 0.1852 | 326103 | 10019 | 2432366 | 35772 | 0 | 0 | 0 | 0 |
|  | 2 | 0.2868 | 169132 | 11801 | 744900 | 39497 | 148980 | 7899 | 137708 | 7302 |
|  | 3 | 0.3599 | 180384 | 17566 | 654600 | 55931 | 490950 | 41948 | 450499 | 38492 |
|  | 4 | 0.5972 | 237242 | 26234 | 576400 | 64023 | 518760 | 57620 | 464852 | 51633 |
|  | 5 | 0.5123 | 64307 | 8736 | 175500 | 25865 | 175500 | 25865 | 158604 | 23375 |
|  | 6 | 0.6187 | 51116 | 7824 | 121000 | 19812 | 121000 | 19812 | 108193 | 17715 |
|  | 7 | 0.5972 | 16505 | 2840 | 40100 | 7308 | 40100 | 7308 | 35933 | 6548 |
|  | 8 | 0.5972 | 20415 | 3481 | 49600 | 9535 | 49600 | 9535 | 44446 | 8544 |
| Total |  |  | 1158146 | 89783 | 8316222 | 258094 | 1544890 | 169987 | 1400234 | 153608 |


| Year: |  | 2008 F multiplie 0.8707 |  |  | Fbar: 0.4545 |  |  | SSB(Jan) | SSNos(ST) | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  | F | CatchNos | Yield | StockNos | Biomass | Nos(Jan) |  |  |  |
|  | 0 | 0.0270 | 81077 | 1118 | 3521756 | 352 | 0 | 0 | 0 | 0 |
|  | 1 | 0.1613 | 298417 | 9168 | 2529347 | 37198 | 0 | 0 | 0 | 0 |
|  | 2 | 0.2497 | 246531 | 17202 | 1225880 | 65000 | 245176 | 13000 | 227467 | 12061 |
|  | 3 | 0.3133 | 112198 | 10926 | 457816 | 39117 | 343362 | 29338 | 316541 | 27046 |
|  | 4 | 0.5200 | 138623 | 15329 | 373965 | 41537 | 336568 | 37384 | 303931 | 33759 |
|  | 5 | 0.4460 | 85331 | 11593 | 259711 | 38275 | 259711 | 38275 | 236267 | 34820 |
|  | 6 | 0.5387 | 32788 | 5019 | 86087 | 14095 | 86087 | 14095 | 77594 | 12705 |
|  | 7 | 0.5200 | 19780 | 3403 | 53360 | 9724 | 53360 | 9724 | 48186 | 8781 |
|  | 8 | 0.5200 | 14982 | 2554 | 40416 | 7770 | 40416 | 7770 | 36497 | 7016 |
| Total |  |  | 1029727 | 76312 | 8548338 | 253070 | 1364681 | 149586 | 1246483 | 136188 |


| Year: |  | 2009 F multiplie 0.8707 |  |  | bar: 0.4545 |  |  | SSB(Jan) | SSNos(ST) | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  | F | CatchNos | Yield | StockNos | Biomass | Nos(Jan) |  |  |  |
|  | 0 | 0.0270 | 81077 | 1118 | 3521756 | 352 | 0 | 0 | 0 | 0 |
|  | 1 | 0.1613 | 299616 | 9205 | 2539505 | 37348 | 0 | 0 | 0 | 0 |
|  | 2 | 0.2497 | 262573 | 18321 | 1305652 | 69230 | 261130 | 13846 | 242269 | 12846 |
|  | 3 | 0.3133 | 191620 | 18661 | 781889 | 66807 | 586417 | 50105 | 540610 | 46192 |
|  | 4 | 0.5200 | 101568 | 11231 | 274002 | 30434 | 246602 | 27391 | 222688 | 24735 |
|  | 5 | 0.4460 | 59807 | 8125 | 182026 | 26826 | 182026 | 26826 | 165595 | 24405 |
|  | 6 | 0.5387 | 51844 | 7936 | 136118 | 22287 | 136118 | 22287 | 122689 | 20088 |
|  | 7 | 0.5200 | 15245 | 2623 | 41126 | 7495 | 41126 | 7495 | 37138 | 6768 |
|  | 8 | 0.5200 | 16920 | 2885 | 45646 | 8775 | 45646 | 8775 | 41219 | 7924 |
| Total |  |  | 1080270 | 80105 | 8827721 | 269554 | 1499065 | 156725 | 1372209 | 142957 |

Table 3.7.5 WESTERN BALTIC HERRING.
Short-term prediction multiple option table, Status quo F.
MFDP version 1a
Run: WBSS GeoMean 5 years
Western Baltic Herring (combined sex; plus group)
Time and date: 14:54 18/03/2007
Fbar age range: 3-6

| 2007 <br> Biomass | SSB | FMult | FBar | Landings |
| :---: | :---: | :---: | :---: | :---: |
| 258094 | 153608 | 1.00 | 0.5220 | 89783 |


| 2008 |  |  |  | 2009 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 253070 | 142291 | 0.00 | 0.0000 | 0 | 354999 | 214596 |
|  | 141933 | 0.05 | 0.0261 | 5159 | 349193 | 209545 |
|  | 141576 | 0.10 | 0.0522 | 10211 | 343510 | 204624 |
|  | 141219 | 0.15 | 0.0783 | 15159 | 337949 | 199831 |
|  | 140864 | 0.20 | 0.1044 | 20006 | 332505 | 195162 |
|  | 140510 | 0.25 | 0.1305 | 24753 | 327175 | 190614 |
|  | 140156 | 0.30 | 0.1566 | 29404 | 321958 | 186183 |
|  | 139804 | 0.35 | 0.1827 | 33960 | 316851 | 181866 |
|  | 139453 | 0.40 | 0.2088 | 38424 | 311850 | 177660 |
|  | 139102 | 0.45 | 0.2349 | 42798 | 306953 | 173561 |
|  | 138752 | 0.50 | 0.2610 | 47084 | 302158 | 169568 |
|  | 138403 | 0.55 | 0.2871 | 51284 | 297463 | 165676 |
|  | 138056 | 0.60 | 0.3132 | 55401 | 292864 | 161883 |
|  | 137709 | 0.65 | 0.3393 | 59435 | 288360 | 158187 |
|  | 137363 | 0.70 | 0.3654 | 63390 | 283949 | 154584 |
|  | 137018 | 0.75 | 0.3915 | 67266 | 279627 | 151073 |
|  | 136674 | 0.80 | 0.4176 | 71066 | 275394 | 147650 |
|  | 136330 | 0.85 | 0.4437 | 74791 | 271247 | 144314 |
|  | 135988 | 0.90 | 0.4698 | 78444 | 267183 | 141061 |
|  | 135646 | 0.95 | 0.4959 | 82026 | 263202 | 137890 |
|  | 135306 | 1.00 | 0.5220 | 85538 | 259301 | 134799 |
|  | 134966 | 1.05 | 0.5481 | 88982 | 255478 | 131784 |
|  | 134627 | 1.10 | 0.5742 | 92360 | 251731 | 128845 |
|  | 134290 | 1.15 | 0.6003 | 95673 | 248059 | 125979 |
|  | 133953 | 1.20 | 0.6264 | 98922 | 244460 | 123184 |
|  | 133617 | 1.25 | 0.6525 | 102110 | 240933 | 120458 |
|  | 133281 | 1.30 | 0.6786 | 105237 | 237474 | 117800 |
|  | 132947 | 1.35 | 0.7047 | 108305 | 234084 | 115207 |
|  | 132614 | 1.40 | 0.7308 | 111316 | 230760 | 112678 |
|  | 132281 | 1.45 | 0.7569 | 114269 | 227501 | 110211 |
|  | 131949 | 1.50 | 0.7830 | 117168 | 224305 | 107805 |
|  | 131619 | 1.55 | 0.8091 | 120013 | 221172 | 105457 |
|  | 131289 | 1.60 | 0.8352 | 122805 | 218099 | 103166 |
|  | 130960 | 1.65 | 0.8613 | 125545 | 215085 | 100931 |
|  | 130631 | 1.70 | 0.8874 | 128235 | 212128 | 98751 |
|  | 130304 | 1.75 | 0.9135 | 130875 | 209229 | 96623 |
|  | 129978 | 1.80 | 0.9396 | 133467 | 206384 | 94547 |
|  | 129652 | 1.85 | 0.9657 | 136012 | 203594 | 92520 |
|  | 129327 | 1.90 | 0.9918 | 138511 | 200857 | 90543 |
|  | 129003 | 1.95 | 1.0179 | 140965 | 198171 | 88612 |
|  | 128680 | 2.00 | 1.0441 | 143374 | 195535 | 86728 |

Input units are thousands and kg - output in tonnes


Figure 3.5.1
WESTERN BALTIC HERRING. Recruitment indices (natural log) adjusted
to year-class, versus time. Estimates of the larval index for 2006 were not available.


Figure 3.6.1 WESTERN BALTIC HERRING.
Proportions of age groups (numbers) in the total catch.


Figure 3.6.2 WESTERN BALTIC HERRING.
Mean weight in the catch ( kg ).


Figure 3.6.3 WESTERN BALTIC HERRING. Estimates of mean F and SSB by ICA runs by individual fleets and catch at age data for 1991-2006.


Figure 3.6.4
WESTERN BALTIC HERRING.
Estimates of mean $F$ and SSB in terminal year by ICA runs by individual fleets and catch at age data for 1991-2006.


Figure 3.6.5 WESTERN BALTIC HERRING. Output from ICA Final Run 2007. Index sum of squares of deviations between model and observations (survey index) as a function of the reference Fin 2006.

Agex 1: Fleet 1b/Dansih Acoustic in Division IIIa+IVaE, ages 2-8+
Agex 2: Fleet 2b/ German Acoustic in SD 22-24, ages 0-5
Agex 3: Fleet 4/IBTS Quarter 3, ages 1-5


Figure 3.6.6

## WESTERN BALTIC HERRING.

Output from ICA Final Run 2007: Stock Summary


Figure 3.6.7
WESTERN BALTIC HERRING.
Output from ICA Final Run 2007.
Separable Model Diagnostics: Log Residual\&Selction Pattern
(Age 0 is still included in the $\log$ residual \& year residuals, although age o was down-weighted (0.1) in the catch)


Figure 3.6.8

WESTERN BALTIC HERRING. ICA Final Run 2007. Log catchability residuals plots.


Figure 3.6.9a WESTERN BALTIC HERRING.
Log Catch vs Age for successive cohorts and their resulting slopes estimates.
CATCH IN NUMBER (CANUM), Ages=1-8


Figure 3.6.9b
WESTERN BALTIC HERRING.
Log Catch vs Age for successive cohorts and their resulting slopes estimates ACOUSTIC SURVEY IN DIV IIIa+IVaE, ages=2-8


Log Catch vs Age for successive cohorts and their resulting slopes estimates ACOUSTIC SURVEY IN SD 22-24 ages=0-5


WESTERN BALTIC HERRING.
Log Catch vs Age for successive cohorts and their resulting slopes estimates IBTS IN KATTEGAT QUARTER 3, ages=1-5


MFYPR version 2a
Run: WBSS GeaMean 5 years_MFYPR
Time and date: 14:56 18/03/2007

| Reference point | F multiplier | Absolute $\mathbf{F}$ |
| :--- | :---: | :---: |
| Fbar(3-6) | 1.0000 | 0.5220 |
| FMax | 1.0117 | 0.5282 |
| F0.1 | 0.4203 | 0.2194 |
| F35\%SPR | 0.4177 | 0.2181 |

MFDP version 1a
Run: WBSS GeoMean 5 years
Western Baltic Herring (combined sex; plus group)
ime and date: 14:54 18/03/2007
Fbar age range: 3-6
Input units are thousands and kg - output in tonnes

Weights in kilograms

Recruitment


Figure 3.8.1 Western Baltic Herring. Various recruitment measures.


Figure 3.8.2 Western Baltic Herring. Histogram of recruitment.

Recruitment


Figure 3.8.3 Western Baltic Herring. QQ plot of recruitment.

## Catch Weight



Figure 3.8.4 Western Baltic Herring. Trends in catch weight by age (1991-2006, ages 0 8+).


1991199219931994199519961997199819992000200120022003200420052006
Year

Figure 3.8.5 Western Baltic Herring. Trend of mean weight in the catch (1991-2006, ages 0 - 8+). Ignoring the 2006 value (indicated by the arrow) leads to the second regression equation.


Figure 3.8.6

[^8]

Figure 3.8.7 Western Baltic Herring. Optimization B. B $_{\mathrm{pa}}=$ lower SSB limit.


Figure 3.8.8
Western Baltic Herring. Optimization C. $\mathrm{B}_{\mathrm{pa}}=$ lower SSB limit.


Figure 3.8.9 Western Baltic Herring. Optimization D. B $_{\mathrm{pa}}=$ lower SSB limit.


Figure 3.8.10 Western Baltic Herring. Optimization E. B pa $=$ lower SSB limit.


Figure 3.9.1
WESTERN BALTIC HERRING.
Historic uncertainty in the Final model fit (ICA assessment). Percentiles 10, 25, 50, 75 and $90 \%$.




Figure 3.9.2 WESTERN BALTIC HERRING: Restrospective Analysis (ICA)


Figure 3.9.3 WESTERN BALTIC HERRING.
Restrospective selection pattern


Figure 3.9.4 WESTERN BALTIC HERRING. SSB estimates from ICA model with separate indices and with indices combined (Final Run 2007).

## 4 Celtic Sea and Division VIIj Herring

### 4.1 The Fishery

The herring fisheries to the south of Ireland in the Celtic Sea and in Division VIIj exploit herring that have a protracted period from September to February. For the purpose of stock assessment and management, these areas have been combined since 1982. The management unit covers all of Divisions VIIg,h,j and $k$ and the southern part of Division VIIa.

### 4.1.1 Advice and management applicable to 2006-2007

## ACFM Advice

The TAC in 2006 was 11050 t , and in 2007 is 9393 t . In 2006, ACFM considered the current level of SSB to be uncertain, but maybe below $\mathrm{B}_{\mathrm{pa}}$ and possibly even below $\mathrm{B}_{\mathrm{lim}}$. There was no short term forecast undertaken but given the risk to the stock indicated by low recruitment, ACFM advised that, fishing should not proceed unless accompanied by a recovery plan. It was proposed that this plan should include the closure of spawning areas and further reduction in catches.

## EU Proposal

In 2006, the EU produced a proposal for the management of this stock, though no explicit recovery plan was developed for the stock. However in 2006, the European Commission produced a proposal on fishing opportunities for 2007. This stock was classified as being "outside safe biological limits". For this stock, the Commission proposed that a TAC for 2007 be set to bring the stock within safe biological limits, but not entailing more than a $\pm 15 \%$ fluctuation in TAC from year to year. However the Commission noted that the TAC should in no case be set at a level that will lead to an increase in F or a decrease in SSB, even if this results in a greater than $15 \%$ reduction in TAC.

## Irish Preliminary Recovery Plan Proposal

Based on the ICES advice, a preliminary rebuilding plan was proposed by Ireland. This plan had the following elements:

1 ) stepwise $15 \%$ reduction in catch year on year
2 ) Indefinite closure of spawning box C (Figure 4.1.1.1a), but with a derogation for vessels below 15 m length allowed to catch a maximum of $8 \%$ of the quota. If the catch of fish $<23 \mathrm{~cm}$ exceeds $50 \%$ then the small scale fishery would be stopped.

This proposal was evaluated by STECF in November 2006. STECF advised that a $15 \%$ reduction in TAC for 2007 results in a high risk that the stock would be below $\mathrm{B}_{\mathrm{pa}}$ in 2008. STECF acknowledged that there were no low risk strategies available at present that would bring the stock to $\mathrm{B}_{\mathrm{pa}}$ in 2008, and that even having no catch represents some risk that SSB would be below $\mathrm{B}_{\mathrm{pa}}$

The Irish plan was not explicitly implemented, though the current TAC represents a $15 \%$ reduction on that for 2006.

Each year, a different "spawning box" is closed for a two week period, at a time corresponding to the nominal peak in spawning activity in that area (Figure 4.1.1.1a). In 2006/2007 box A was closed and in 2007/2008 Box B will be closed.

## Local Irish Management Plan

A committee was established to manage the Irish fishery for this stock, on a local basis. This committee, therefore, has responsibility for management of the entire fishery for this stock at present. The committee stated its intention to follow the following objectives:

- To build the stock to a level whereby it can sustain annual catches of around 20,000 t.
- In the event of the stock falling below the level at which these catches can be sustained the Committee would take appropriate rebuilding measures.
- To introduce measures to prevent landings of small and juvenile herring including closed areas, and or appropriate time closures.
- To ensure that all landings of herring should contain at least $50 \%$ of individual fish above 23 cm .
- To maintain and if necessary expand, the spawning box closures in time and area.
- To ensure that adequate scientific resources are available to assess the state of the stock.
- To participate in the collection of data and to play an active part in the stock assessment procedure.


### 4.1.2 The fishery in 2006/2007

The landings in this fishery since 1958 are shown in Figure 4.1.2.1.
In 2006-2007, 40 vessels took part in the Irish fishery. These are categorised as follows:

- 7 Pelagic RSW trawlers from 45 m to 27 m
- 6 Polyvalent RSW trawlers 24 m to 33 m
- 8 Polyvalent tank hold trawlers from 19 m to 26 m
- 19 Polyvalent dryhold trawlers from 10 m to 27 m

The fishery took place in the third, fourth and first quarter. Most vessels under 20 m reported landings of less than $100 t$ for the assessment period while a number of RSW vessels reported landings greater than 1000 t . In addition small incidental landings, typically less than a tonne were reported by a number of other vessels. The term "Polyvalent" refers to a segment of the Irish fleet, licensed to catch a variety of species, under Irish law. "Pelagic" segment vessels are confined to fishing pelagic species.

The third quarter fishery took place between the Kinsale gas fields and Labadie Bank, being prosecuted by Polyvalent RSW and Polyvalent tank hold boats, in late August (Figure 4.1.1.1b). The fourth quarter fishery lasted from the third week of November to the $20^{\text {th }}$ December., being largely restricted to VIIa south. In the first quarter of 2007 the fishery was constrained by bad weather and the loss of two vessels, being only open from the $2^{\text {nd }}$ to the $19^{\text {th }}$ of January. The fishery closed early due to search and rescue efforts. When the fishery was open landings came from the spawning grounds in VIIg (Figure 4.1.1.1b). Overall about 900 t were taken from around Waterford harbour, by smaller vessels.

The Irish quota is managed by allocating individual quotas to vessels on a weekly basis. Participation in the fishery is restricted to licensed vessels. The licensing requirements have been changed in recent years. Previously, vessels had to participate in the fishery each season to maintain their licence. Now this requirement has been lifted. This has been one of the contributing factors to the reduction in number of vessels participating in the fishery in recent seasons. The efficiency of these vessels has improved, however. Fishing is restricted to the period Monday to Friday each week, and vessels must apply a week in advance before they are allowed to fish in the following week.

### 4.1.3 The catches in $2006 / 2007$

The estimated national catches from 1988-2006 for the combined areas by year and by season (1 April-31 March) are given in Table 4.1.3.1 and Table 4.1.3.2 respectively. The catch, taken during the 2006/2007 season was under 7000 t having declined from almost 9500 t in the previous season (Figure 4.1.2.1.).

There are no estimates of discards for this fishery. Anecdotal reports from fishermen suggest that discarding is not a feature of this fishery at present.

### 4.2 Biological composition of the catch

### 4.2.1 Catches in numbers-at-age

Catch numbers at age are available for the period 1958/1959 to 2006/2007. These data include discards, until 1997, and afterwards no discard estimates have been available (Table 4.2.1.1). In 2006/2007, there was a strong dominance of 2-ringers (2003/2004 year class). This cohort was strong the previous season also. These two ringers are stronger than any year class since 2000/2001. The 4-ringers (2001/2002 year class) were weak as in previous seasons (Table 4.2.1.1). It is important to note that the weakness of 4-ringers tends to inflate the proportions of the other important age groups in the catches. However the yearly mean standardised plots (Figure 4.2.1.1.) shows that 2-ringers have been the dominant age in catches in general throughout the series. It can also be seen that although younger fish always predominated, there is a marked truncation of the age profile since the 1980s. In most recent years this effect seems to have accentuated.

The overall proportions at age were similar in all sampled metiers (division*quarter), see Figure 4.2.1.2. The quarter 3 fishery in VIIg produced greater proportions of 5 ringers than elsewhere. As usual, VIIj catches had a different age distribution to the other areas. Table 4.2.1.2 shows that there were very small fish recorded in VIIaS and that largest fish were caught in VIIg and VIIj.

### 4.2.2 Movements of fish

## Juveniles

It was shown that fish of Celtic Sea origin are present in the western Irish Sea, and then return south as 1- and 2-ringers (Molloy et al. 1993). This was endorsed by Brophy and Danilowicz, (2002) and confirmed by the WESTHER project (Hatfield et al., 2007 WD). Historic larval survey data from the 1980s (Stock Annex) show that autumn and winter spawning was taking place in the eastern Celtic Sea (ICES, 2006). Thus juveniles of Celtic Sea origin present in the Irish Sea cannot be distinguished on the basis of spawner type alone. This creates problems with some techniques that could be used to separate Irish sea catches by spawning origin. Further work is required on the effect of this juvenile mixing on the assessment of the Celtic Sea stock.

## Adults

The quarter 3 fishery targets offshore feeding aggregations of herring in VIIg. It is not known where these fish spawn. However a combination of positional data from the commercial fishery (August) and acoustic surveys (September to December) in this area displays a seasonally progressive movement towards the traditional spawning grounds inshore (Stock Annex). The WESTHER project (Hatfield et al. 2007 WD), shows that VIIj spawners were found feeding in VIIg. This supports the current assessment and management units. Even though VIIj fish display a different age profile and recruitment dynamic, they feed in the same region as fish that return to spawn in VIIg and VIIaS. It is not known if fish from VIIe also feed in this area (Stock Annex).

### 4.2.3 Quality of catch and biological data

Biological sampling of the catches throughout the region was comprehensive, except in VIIj. The spasmodic nature of the fishery in that area makes sample acquisition difficult, (Table 4.2.3.1). Under the Data Collection Programme the sampling of this stock is well above that required by the Minimum Programme (Section 1.5). An analysis of precision of Irish herring catch at age data shows excellent quality data (CVs $<7 \%$ ) for the main ages in the catches from this stock.

The quality of catch data has varied over time. Table 4.2.3.2 presents a rudimentary history of the Irish fishery since 1958. The quality of landings data has improved in most recent years, particularly since 2004, when a low tolerance for water in catches was introduced. The change in water content, changes in control and the demise of the roe fishery all point to better data quality. These factors may bias the data in the respective periods, and such biases need to be considered when examining long term stock dynamics.

Discarding was a major feature of the fishery during from 1983 to 1997, when the fishery sought fish of a particular roe quality, discarding early stage, spent and young fish. Though discarding is thought to be lower in subsequent years, the tight quota situation coupled with market requirements are known to lead to some discarding, particularly of smaller fish.

In 1991 the working group revised the catches, to account for possible underreporting. This was done by calculating the catch associated with the roe production in Ireland (ICES, 1991). In 1992, this procedure was reversed because of concerns that the roe production data used included material from stocks other than this one (ICES, 1992). The 1991 revision scaled up the working group catch in the period 1983/1984 to 1990/1991 by between 20 and $34 \%$. Though this approach was not without its problems it could be an indicator of the scale of unaccounted landings in the roe fishery period (1983-1997). The realised catch in this period may have even higher, partly at least due to further unaccounted discarding. There is no information on misreporting in this fishery in recent years, but it is thought to have decreased.

### 4.3 Fishery Independent Information

### 4.3.1 Acoustic Surveys

Acoustic surveys of this stock have been carried out since 1990, with the exception of 1997. Up until 1996, two acoustic surveys were carried out annually. In 1997 there was no research vessel available to do the survey. Since 1998, usually only one winter survey was conducted (Table 4.3.1.1). The acoustic series was revised (ICES, 2006; Stock Annex). This series dates from 1995 and is presented in Table 4.3.1.2.

The acoustic survey of the 2006/2007 season was carried out in October 2006, on the Celtic Explorer, and was the most comprehensive in the current time series. The survey track began at the northern boundary of VIIj, covering the SW bays in zig-zags and parallel transects. The main broad scale survey in VIIg and VIIaS adapted a parallel transect design transect spacing of 4 nmi in areas of low historic abundance and 2 nmi spacing in areas of high historic abundance. The survey extended 78 nmi offshore (Figure 4.3.1.1a) to include the offshore trawl grounds. A detailed survey of autumn spawning grounds was undertaken after the main broad scale survey was complete. Spawning grounds were surveyed working in an east to west progression using either 1nmi parallel transects for larger grounds or detection using the vessels sonar for discreet spawning beds In total the combined survey transect length was 2901 nmi.

The south western region contributed little herring to the overall estimate (Figure 4.3.1.1b). This area has previously been the dominant area for juvenile herring. Small amounts of herring were encountered during the broad scale survey, appearing in mixed schools over a
relatively small geographical area. Traditional inshore herring spawning areas in VIIg and VIIaS were found to contain the greatest herring biomass, and made up 77\% of the TSB.

The age structured index of biomass and catch numbers from acoustic surveys in this area, is shown in Table 4.3.1.2. In 2006/2007 the SSB estimate was 35974 t .

The percentage age composition in the survey and the commercial fishery are compared in Figure 4.3.1.2. The survey displayed the same age distribution as the commercial fishery, both showing a strong predominance of 2-ringers and the very low abundance of 1 and 4 -ringers. The survey showed somewhat lower numbers of 5 and 6 -ringers and did not pick up any older fish.

### 4.3.2 Other surveys

Previous working groups examined the utility of other surveys as tuning indices. These surveys included the EVHOE quarter 4 IBTS survey, the Irish Groundfish quarter 4 IBTS survey and the UK (E\&W) quarter 1 Portuguese high headline survey (ICES, 2005; 2006, Stock Annex). None of these surveys were particularly useful, all having noisy data and strong year effects. However the Irish survey did pick up the weakness of the 2001/2002 year class.

Existing surveys that may have utility for tuning the assessment are the DARDNI Groundfish Survey of the Irish Sea, Northern Ireland and the UK quarter 4 western GFS. It is known that juveniles from the Celtic Sea are present in the Irish Sea. If it is possible to distinguish these fish from native Irish Sea herring, in the AFBI survey then this survey could offer potential for a recruit index for Celtic Sea herring.

Given that this fishery is dependent on recruiting year classes, a recruit survey would be very helpful.

### 4.4 Mean weights-at-age and maturity-at-age

The mean weights in the catch over time are presented in Figure 4.4.1, with stock weights displayed in Figure 4.4.2. There has been an overall downward trend in mean weights at age since the mid-1980's. The values for 2006/2007 for the important age groups are among the lowest in the series. This trend in mean weights at age is similar to those seen in VIaN, the Irish Sea and to a lesser extent, the North Sea.

Mean weights in the stock at spawning time were calculated from biological samples, excepting VIIg quarter 3. The numbers of fish sampled for 7 to 9 ring were low.

The current assessment considers $50 \%$ of 1-ringers to be mature, but the percentage is higher, at least in commercial catches. A new project to develop maturity ogives for this stock from catch and survey data started in 2006 (Lynch, in prep). This project will also examine long term changes in biological parameters. It is known that more than $50 \%$ of 1-ringers are mature in some years.

### 4.5 Recruitment

At present there are no recruitment estimates for this stock that can be used for predictive purposes. The 2003 recruitment was estimated as weak in the 2004 assessments, and appears to be the weakest in the series. There is little information in the assessment on the strength of recruitment in any year, because these 1-ringers are poorly represented in the catches.

A rudimentary analysis of recruitment, in comparison with long term geometric mean, over the time series suggests a historical pattern as follows:

1958-1972: Mostly above average, strong year class roughly every 5th year
1973-1980: Low recruitment, all below average.
1981-1995: Mostly above average, strong year class roughly every 5th year
1996-recent: Around average, with one very poor year class.
Variance around estimated recruitment is broadly similar in all periods.
The possibility that the stock has entered a new period of low productivity will be investigated further.

### 4.6 Assessment

This stock is scheduled for benchmark assessment in 2007. The last time the assessment of this stock was accepted by ACFM as a basis for management advice was 2001. In 2006, the working group continued to conduct exploratory assessments and no final assessment was put forward. The most important information considered by the 2005 working group was weakness of the 2001 year class.

### 4.6.1 Data exploration

Data exploration consisted of examining a number of features of the basic data. These analyses included log catch ratios, cohort catch curves in survey and catch at age series.

Log catch ratios were constructed for the time series of catch at age data, as follows:

$$
\log [C(a, y) / C(a+1, y+1)]
$$

These are presented in Figure 4.6.1.1. It can be seen that 1-ringers, and the oldest ages, have a noisy signal, being poorly represented in the catches. Overall there is a trend towards greater mortality in recent years. The increased mortality visible in the older ages corresponds with the truncation in oldest ages in the catch at age profile (Figure 4.2.1.1.). There was an increase in ratios in 1998, that seems quite abrupt. It can also be seen that the gross mortality signal was low in 2002, corresponding to the big decrease in catch in that year. The signal increased again in 2003, concomitant with increasing catch.

Cohort catch curves, showing raw total mortality Z per year class, were constructed for each year class in the catch at age data and for year classes in the acoustic time series where enough data were available. These are displayed in Figure 4.6.1.2, and the Z estimated over 2-7 ringer is shown in Table 4.6.1.1.. Total mortality was low for cohorts 1956 to 1964 (Figure 4.6.1.3.). Cohorts in the late 1960s seem to display higher Z, but those from 1975 to 1982 displayed the highest Z ( 0.6 to 1.1). The most recent year classes for which enough observations are available (1991-1997) show higher Z again, in the range about 0.6 to 1.0 . There is a marked secondary peak in all the cohorts, corresponding to the 2001/2002 fishing season. It is considered that this corresponds to the closure of Spawning Box C in that year, which shifted exploitation to the western part of the Celtic Sea where older fish are usually caught.

Cohort catch curves are shown for the catch at age data (Figure 4.6.1.4) and for the acoustic survey (Figure 4.6.1.5). The same patterns in raw mortality are visible, but the Zs from the acoustic survey are somewhat higher than those from the commercial data. This may be explained as differing catchability between the two, and it should be noted when interpreting the assessment results below.

### 4.6.2 Exploratory Assessments

In 2007, exploratory assessments were conducted to investigate a number of issues.

- Validity of separable assumption
- Shape of the exploitation pattern
- Choice of the separable period.
- WINBUGS Bayesian analysis
- Use of simple stage structured model

All these were performed on the revised acoustic series, that was described in ICES (2006).

## ICA and XSA assessments

In each case, the same procedure as previous years was used to deal with the assumption that $50 \%$ of 1-ringers are mature. Therefore recruitment at 1-ring was replaced with geometric mean (1958-2004), and the SSB was recalculated based on the stock weights, maturity ogive and population numbers in the final year.

The ICA base case exploratory run was set using similar settings as in previous years. The separable period was set at 6 years (Table 4.6.2.1). This was considered valid because of the $a$ priori assumption that the exploitation pattern shifted in 2001. This shift is visible in the cohort catch curves (Section 4.6.1 and Figure 4.6.1.2). After 2001, the summer feeding ground fishery also developed, and this is captured in this separable period. Shorter and longer separable periods were also tested. The residual pattern for the base case and the run with 8 year separable period are shown in Figure 4.6.2.1. There was no appreciable improvement in model fit when extending the separable period beyond the base case 6 years. Shorter separable periods were not considered informative based on the a priori assumption nor were they considered sufficiently long to achieve a good model fit. There were no differences in the survey index residuals (Figure 4.6.2.2).

It seems that exploitation pattern may not be flat topped. This is supported by the 2005 observation, from bycatch samples from demersal seine net vessels, that older herring were present but not well represented in the catches. Selection at oldest true age tested in ICA by screening over the range 1 to 0.7 , relative to 3-ringer. Though the dome shaped pattern may be more appropriate, it did not improve the residual pattern (Figure 4.6.2.1).

Assessments of this stock have used 9-ring as the plus group. However, in recent years, few fish older than 6 -ring occur in catches. Therefore a reduction in the plus group to 7 -ring was investigated. Again, this did not improve the residual pattern (Figure 4.6.2.1).

A comparison XSA and ICA was conducted, because there were concerns about the validity of the separable assumption. The XSA was run with low shrinkage, whilst the ICA base case was run with no shrinkage. There were no diagnostics that could be used to choose one model over the other. The catchability residual pattern from the XSA shows strong year effects (Figure 4.6.2.3). However, HAWG considered that ICA to be more useful for this stock. This is because XSA was considered to be more influenced by the older ages, due to the weighting procedures. These older ages are not present in catches. On the other hand, ICA is more influenced by younger ages, and was considered more useful in this stock where younger ages are dominant. It was also considered that ICA is more robust to noise in the survey data.

None of these exploratory cohort analyses are better than any other, and all have poor residual patterns. This shows that there are inconsistencies with the input data, rather than the models themselves. The input data are simply very noisy and this cannot be dealt with by choosing one model over another. Therefore it is necessary to evaluate the information available, in the light of the model assumptions and also based on the exploration of the raw data.

Comparison of the recruitment, SSB and F from the trial assessment runs using XSA and ICA are presented in Figure 4.6.2.4. All runs of ICA show very similar trends over time and in the
most recent years. Thus, the recent stock history is reasonably clear, whatever model or setting is used.

Changing the separable period did not change stock perception appreciably. Changing the selection pattern did not change trends in the outputs over time, but it did introduce a positive scaling factor on SSB over time. Reducing the plus group to 7-ring, did not change perception of the SSB or recruitment. However, it should be noted that the mean F trend is different, simply because it is calculated over 2-5 ring and not 2-7. This was to avoid including the plus group in the mean F calculation.

XSA displays the same trends in SSB and recruitment as the ICA base case. However, F in the final three years is higher from the XSA run, relative to ICA (Figure 4.6.2.4). Both models show very high F in 2003. This trend was also shown in the log catch ratio (Figure 4.2.1.1). XSA follows the trend in the log catch ratios, that continue to be high in the main ages (2-5) after 2003.

## Catch Survey Analysis

An exercise using Catch Survey Analysis (CSA, Mesnil) was undertaken to explore the results of using a comparatively simple model to assess the state of the stock. The population dynamics is described by the following model:

$$
\begin{equation*}
N_{y+1}=\left(N_{y}+R_{y}\right) e^{-M}-C_{y} e^{-M(1-\tau)} \tag{1}
\end{equation*}
$$

where:
y : time step, typically annual. Years may be defined either on a calendar basis or as the interval between regular surveys. The year range is [1, Y].

Ny : population size, in number, of fully recruited animals at start of year y ;
Ry : population size, in number, of recruits at start of year y;
Cy : catch in number during year y (known);
My : instantaneous rate of natural mortality (equal for both stages, assumed);
$\tau$ : fraction of the year when the catch is taken, e.g. 0 if the fishing season is early in the year, or 0.5 if the catch is taken midway through the year or, by resemblance with Pope's (1972) cohort approximation, evenly over the year.

Estimating the time series of Ny and Ry given the catches is the basic task of any assessment but, as with other methods, this requires additional information in the form of relative indices ny and ry of abundance for each stage, typically from surveys, which are assumed to be proportional to absolute population sizes Ny and Ry. The indices are deemed to be measured with some (log-normal) observation error:

| $\mathrm{N}_{\mathrm{y}}$ | $=$ | $\mathrm{q}_{\mathrm{n}} \mathrm{N}_{\mathrm{y}} \exp \left(\mathrm{n}_{\mathrm{y}}\right) ;$ | y | $=$ | 1, |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Ry | $=$ | $\mathrm{q}_{\mathrm{r}} \mathrm{R}_{\mathrm{y}} \exp \left(\delta_{\mathrm{y}}\right) ;$ | y | $=$ | 1 |

where:
qn and qr : catchability coefficients of fully-recruited and recruits, respectively, in the survey, supposed to be constant with time;
$\eta$ and $\delta:$ normally distributed random variables.
A constraint must be imposed whereby the survey catchability of the recruits is some fraction s of that of the fully-recruited:
$\mathrm{S} \quad=\quad \mathrm{q}_{\mathrm{r}} / \mathrm{q}_{\mathrm{n}}$
Data used (Table 4.6.2.2) were the catch numbers from 1995 to date, the stock weights at ages 2 and $3+$, the acoustic survey series of the numbers of the 2 year-old group (taken as recruits) and the numbers of age 3 and older, fully recruited ages (Table 4.6.2.2). Natural mortality (M) was estimated as a weighted mean based on the stock numbers at age $2+$ matrix and equal to 0.21 .

The model was run assuming that catches take place in the middle of the year. As it appears that the catchability ratio $s$ cannot be estimated together with other parameters, the model was run iteratively for a set of values of $s$. A minimum in the sum of squares was found for $s=0.6$. Results presented correspond to the parameter scenario yielding the best fit.

The model estimated total biomass for the period 1995 - 2006 is shown in Figure 4.6.2.5. The SSB trajectory as estimated by ICA base case run shows a similar biomass level and declining trend. However, CSA is suggesting a further decline in biomass in the recent couple of years. Comparison between the estimated numbers at age scaled by their corresponding catchabilities and the survey indices suggests a difficulty in fitting the data given the number of missing observations (Figure 4.6.2.5).

It should be noted that CSA is not directly comparable with VPA models, and that the age group that is termed "recruits" at 2-ring is not the same as recruitment (1-ring) in the VPA analysis. It should also be noted that 2-ring is the first fully selected age in the acoustic series and therefore the estimated ratio of catchability (s) between 2- and 3+ in CSA may not be appropriate. However, this effect could be associated with $M$, which is expected to be correlated with the estimation of S.

However the CSA run does show the same overall downward trends as the VPAs. The overall trend, for a fall in biomass since 1997 is supported by CSA.

## Exploration of Celtic Sea assessment with WINBUGS

To explore the Celtic Sea data set further and to explicitly incorporate error in the assessment a Bayesian model was set up in WINBUGS (see section 1.6). The code is included in Table 4.6.2.3. The model has been set up as a statistical catch at age approach for the period where survey data is available (1995 to 2006) The early part of the series is derived from a VPA with starting numbers from 1995. The catch at age is fitted with a logistic selection function that can change from year to year. The example shown has allowed only a slow change in selection.

The results of an assessment run are illustrated in Figure 4.6.2.6 and indicate considerable uncertainty in SSB in the terminal year with signs that $\mathrm{F}_{2-5}$ has reduced in recent years. The flexible selection pattern shows a twist in selection over years, with a steeper logistic curve in later years compared to earlier years. The scatter plots of modelled catch against observed catch and modelled survey against observed survey (Figure 4.6.2.7) show that the catch fits more closely in the model than the survey.

This model exploration is very preliminary, there has been no evaluation of the influence of priors, though they are thought to be uninformative, and only a limited range of flexible selection has been tested.

## Estimates of uncertainty

This approach does show the large uncertainty around F since the mid 1990s. This is also displayed in the uncertainty analaysis from the ICA base case (Figure 4.6.2.6). Although mean F is calculated over different age ranges between WINBUGS and ICA, it is clear that median F estimates for 2002 were lower than other recent estimates. Also, F in subsequent
years may be declining, though there is very large uncertainty. Most recent SSB is very poorly estimated.

The uncertainty analysis, based on bootstrapping of the ICA base case run , shows that uncertainty began to increase very markedly from 1995 onwards. This is most likely explained by inconsistencies between cohort representation in the survey and the commercial catch at age data. These uncertainty estimates are based on bootstrapping of the variance covariance matrix. The lack of uncertainty in earlier years is because this is a traditional VPA and not because there was less uncertainty in that period.

## Exploration of recruitment dynamics

The trend in recruitment anomalies ( Ra ) of Celtic Sea herring was analysed in order to conduct a rudimentary analysis of the effect of environmental conditions (SST and NAO winter index) on recruitment dynamic. The approach used here has been advocated by recent works of Stige et al. (2006), and Cardinale and Hjelm (2006):

The logarithm of the ratio between the annual numbers of recruits and the SSB is considered a robust proxy for the recruitment success (Rs) of a stock and ideal for disentangling environmental to adult biomass effect on recruitment (Beverton 2002, Cardinale and Hjelm 2006, Stige et al. 2006). According to the models (Ricker 1954) and Beverton and Holt (1995), this ratio is a linear function of SSB:
(1) $\ln (R / S S B)=a+b S S B$
where other effects can be modelled as predictors in (1).
Therefore, the variability around the relationship between Rs and SSB can be considered as a proxy for recruitment anomalies ( Ra ) and assumed to be partially determined by the stochasticity in the physical environment (Beverton 2002) as well as other biotic factors. The rationale behind this approach is mathematically formalized in the classical Ricker (1954) or Beverton and Holt (1995) recruitment functions where the number of recruits in a fish species is generally related to egg number or SSB, a proxy of egg production (Myers and Barrowman 1996) and these functions are in turn based on sound ecological mechanisms (e.g. cannibalism and predation).

Thus, after fitting (1), the standardized residuals $\left(R_{a}\right)$ for this stock were estimated (Figure 4.6.2.9). There was a consistent period of low recruitment in the 1970s, and a consistent period of high recruitment in the early 1980s. The current period is characterised by more variable recruitment around the mean, and this may be more like the situation in the 1960s, though at that time there was a more positive tendency in the anomalies. Thus, the current period is either similar to the 1960s, but with less extremes, or else it is a less fluctuating new state.

Beverton and Holt (1995), Ricker (1954) and Hockey Stick models were applied to the observed SSB-R relationship and compared with the standardized residuals against Ra. Residual patterns were similar in all cases (Figure 4.6.2.9).

Time series of SST for the Celtic Sea were obtained from Rayner at al (2003). SST data for peak spawning (December), in the winter (January-February) and during the entire spawning season (September to March) were available. There was no discernible relationship between the residual patterns with SST or the Hurrel Index of NAO over the time series.

The analysis of productivity of herring stocks (Section 1.8.3) shows that numbers of recruits per spawner has remained quite constant throughout the series. This is in contrast to Irish Sea and North Sea herring that both displayed larger changes in productivity (Figure 1.8.3.2c). The results of this analysis suggest that median long term yield in this stock would not be higher than 17000 t , assuming that the stock recruit relationships used are truly representing
the reality. It should be pointed out that this is a preliminary analysis and that stochastic simulations should be carried out to investigate the risk associated with such a catch policy. However it can be concluded that the management committee objective of building the stock to a level concomitant with catches of 20000 to per year is not achievable. This accords with yield per recruit calculations conducted by Burd and Bracken (1966) that suggested long-term yields of between 12000 and 15000 .

## Conclusions

The stock history was divided into a number of time periods to aid a long term understanding of the stock dynamics (Table 4.6.2.4). It is clear that growth rate has changed over time. Mean length and mean weight at age have declined by about $15 \%$ and $30 \%$ respectively since the late 1970s. Fish are shorter and lighter at age now than at any time in the series. There is a clear trend for fewer older fish in the catch than in earlier times. Only the cohorts from before the stock collapsed and a few from the late 1980s contributed many older fish appear in the catches. Raw mortality signals, from cohort catch curves suggest that some of the recent year classes have displayed a higher total mortality. The period of sustained below geometric mean recruitment only occurred once in the stock history, from roughly 1973 to 1980. The periods before and after that were more variable though producing high recruitments. The current recruitment pattern seems to oscillate around geometric mean. The 2001 year class was very weak and the 2003 year class seems very strong. If the stock is to recover, this recent cohort should be allowed to contribute to the stock in the next years.

Possible bias may exist in the data in the above history, due to unaccounted mortality. In particular, during the roe fishery from 1983 to 1997, realised catches would have been higher and though the actual catch in more recent periods is not without error, but accuracy is considered to have improved a lot.

Though no final assessment run was conducted by the working group, it is clear that there is an overall downward trend in SSB in recent years. There is a retrospective bias, so that the final SSB is often overestimated. There was an upturn in SSB and a downturn in F in 2002, but the trend was continued after that. F was high in 2000 and again in 2003. However the uncertainty in these F estimates are very large. The 2001 year class was the lowest in the series. Now that 4 observations of this cohort are available in the data, this confirms earlier perceptions of the weakness of this year class. Uncertainty in estimated recruitment was high from 1997 to 2002, but since the uncertainty is much lower, lending belief to estimates of the fully recruited year classes.

The systemic problem in this assessment remains: that the incoming year class is not well estimated, in the absence of a recruit index. The 2003 year class seems very strong however. In the absence of recruit survey information, it is necessary to deal with the data available and to use this to produce management advice. Though current stock size is unknown and F poorly estimated, it is still possible to put forward some information basis of advice. Trends in mortality have increased in recent years, SSB has declined since the mid 1990s, condition of fish has declined. Though the absolute SSB level is unknown it probably below Bpa and possibly below Blim.

The stock has not displayed strong changes in productivity (recruits per spawner) over time. Nor is there any evidence of an environmental driven regime shift, in contrast to the North and Irish Sea. However recruitment dynamics are somewhat different to previous periods, with less variability.

### 4.7 Short term projections

There was no final assessment and consequently no short term projections.

### 4.8 Medium term projections

A yield per recruit was conducted in 2007, see Figure 4.8.1. $\mathrm{F}_{0.1}$ and F at $35 \%$ of spawners per recruit are both estimated as 0.19 . $\mathrm{F}_{\text {med }}$ is estimated at 0.28 .

### 4.9 Precautionary and yield based reference points

Biological reference points were discussed in detail at the working groups in the late 1990s (ICES 1998; ICES 1999). A summary of this discussion was presented in the 2002 HAWG report (ICES 2002). The SGPRP (ICES, 2003) reviewed the methodology for the calculation of biological reference points, and applied a segmented regression to the stock and recruit data from the 2002 HAWG assessment, finding a breakpoint at 61306 t . This change point was considered very high. HAWG decided that the first priority for this stock should be to achieve a stable assessment and that once this was done the reference points would be reinvestigated.

There is still considerable instability in the assessment, so there is no basis for a revision of reference points at this point. $\mathbf{B}_{\mathrm{pa}}$ is currently at 44000 t (low probability of low recruitment) and $\mathbf{B}_{\text {lim }}$ at $26000 \mathrm{t}\left(\mathbf{B}_{\text {loss }}\right)$ for this stock $\mathbf{F}_{\text {ра }}$ and $\mathbf{F}_{\text {lim }}$ are not defined.

A recent management strategy simulation (STECF, 2006; Kelly and Campbell, 2006) estimate the break point in a hockey stick stock recruit model to be around $44,000 \mathrm{t}$. This suggested that the definition of PA points for this stock were unsuitable e.g. Bpa should be Blim, and that an HCR should be devised with a trigger biomass far enough above Blim to prevent recruitment impairment given assessment uncertainty. It is important to differentiate between a breakpoint for the purposes of harvest strategy development, and precautionary reference points for the purposes of advice.

### 4.10 Quality of the Assessment

No assessment was conducted, and the basic data and exploratory analyses are presented above. However to investigate possible retrospective patterns in the past few years, a retrospective analysis was created using the ICA base case. This is presented in Figure 4.10.1. There is no appreciable retrospective pattern in F. The ICA base case tends to over estimate recruitment and SSB in the terminal year, though it should be noted that these estimates are always adjusted by using geometric mean for 1-ringers.

Though no assessment is put forward all the trial assessments show the same downward trend, that the stock has declined since the mid 1990s, to a lower level.

### 4.11 Management Considerations

Though this was a benchmark assessment, no final assessment was conducted, despite a lot of exploratory work. However, there are certain pieces of information that can be obtained from the available data, that can be used to frame management considerations.

Recent recruitment has fluctuated around the mean, with a poor and an evidently good year class. SSB has shown an overall downward trend since 1994. The truncation of age groups in the catches, the decreasing SSB, increasing F in recent years and the poor recent recruitment (2001) are causes for concern. There seems to be one very good recruitment (2003) must be utilised to help rebuild the stock. The stock is probably at as low a level as when it previously collapsed. It is not possible to estimate current stock size, with precisions. Such poor precision should not be the main consideration for management. Instead, management should try to reverse the overall trend, and bring the stock back to a higher overall level.

HAWG advocates a structured approach to the development of a rebuilding plan for this stock. Much useful work on this subject has been presented in the recent SGMAS report, and it is
proposed to proceed along these lines. In particular attention should be placed on the knowledge base available. The most important issues are considered to be:

- Stock characteristics, historically and in the present period
- Decreases over time in growth and condition of the fish
- There is a time lag between the estimation of incoming year classes in the assessment, and their appearance in the fishery.
- Apparent lack of changes in environmental regime
- Development of new management objectives, since the current aspiration now appear unattainable.
- Interact with stakeholders concerns, fleet segments and processing sector issues, management issues.

It is hoped to develop a rebuilding plan in the next year. This will be developed with stakeholders and following best practice as outlined in SGMAS.

Table 4.1.3.1. Celtic Sea and Division VIIj herring. Landings by quota year (t), 1988-2006. (Data provided by Working Group members.) These figures may not in all cases correspond to the official statistics and cannot be used for management purposes.

| Year | France | Germany | Ireland | Netherlands | U.K. | Unallocated | Discards | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
| 1989 | + | - | 16,000 | 1,900 | - | 1,300 | 3,500 | 22,700 |
| 1990 | + | - | 15,800 | 1,000 | 200 | 700 | 2,500 | 20,200 |
| 1991 | + | 100 | 19,400 | 1,600 | - | 600 | 1,900 | 23,600 |
| 1992 | 500 | - | 18,000 | 100 | + | 2,300 | 2,100 | 23,000 |
| 1993 | - | - | 19,000 | 1,300 | + | $-1,100$ | 1,900 | 21,100 |
| 1994 | + | 200 | 17,400 | 1,300 | + | $-1,500$ | 1,700 | 19,100 |
| 1995 | 200 | 200 | 18,000 | 100 | + | -200 | 700 | 19,000 |
| 1996 | 1,000 | 0 | 18,600 | 1,000 | - | $-1,800$ | 3,000 | 21,800 |
| 1997 | 1,300 | 0 | 18,000 | 1,400 | - | $-2,600$ | 700 | 18,800 |
| 1998 | + | - | 19,300 | 1,200 | - | -200 | - | 20,300 |
| 1999 |  | 200 | 17,900 | 1300 | + | -1300 | - | 18,100 |
| 2000 | 1,359 | 219 | 17,729 | - | - | -1578 | - | 17,729 |
| 2001 | 734 | - | 10,550 | 257 | - | -991 | - | 10,550 |
| 2002 | 800 | - | 10,875 | 692 | 14 | $-1,506$ | - | 10,875 |
| 2003 | 801 | 41 | 11,024 | - | - | -801 | - | 11,065 |
| 2004 | 821 | 150 | 8452 | 799 | - | -1770 | - | 8,452 |
| 2005 | - | - | 8,530 | 518 | 5 | -523 | - | 8,530 |
| 2006 |  |  |  |  |  |  |  | -617 |

Table 4.1.3.2. Celtic Sea \& Division VIIj herring landings (t) by assessment year (1st April-31 ${ }^{\text {st }}$ March) 1988/1989-2006/2007. (Data provided by Working Group members.) These figures may not in all cases correspond to the official statistics and cannot be used for management purposes.

| Year | France | Germany | Ireland | Netherlands | U.K. | Unallocated | Discards | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
| $1989 / 1990$ | + | - | 15,000 | 1,900 | - | 2,600 | 3,600 | 23,100 |
| $1990 / 1991$ | + | - | 15,000 | 1,000 | 200 | 700 | 1,700 | 18,600 |
| $1991 / 1992$ | 500 | 100 | 21,400 | 1,600 | - | -100 | 2,100 | 25,600 |
| $1992 / 1993$ | - | - | 18,000 | 1,300 | - | -100 | 2,000 | 21,200 |
| $1993 / 1994$ | - | - | 16,600 | 1,300 | + | $-1,100$ | 1,800 | 18,600 |
| $1994 / 1995$ | + | 200 | 17,400 | 1,300 | + | $-1,500$ | 1,900 | 19,300 |
| $1995 / 1996$ | 200 | 200 | 20,000 | 100 | + | -200 | 3,000 | 23,300 |
| $1996 / 1997$ | 1,000 | - | 17,900 | 1,000 | - | $-1,800$ | 750 | 18,800 |
| $1997 / 1998$ | 1,300 | - | 19,900 | 1,400 | - | -2100 | - | 20,500 |
| $1998 / 1999$ | + | - | 17,700 | 1,200 | - | -700 | - | 18,200 |
| $1999 / 2000$ |  | 200 | 18,300 | 1300 | + | -1300 | - | 18,500 |
| $2000 / 2001$ | 573 | 228 | 16,962 | 44 | 1 | -617 | - | 17,191 |
| $2001 / 2002$ | - | - | 15,236 | - | - | - | - | 15,236 |
| $2002 / 2003$ | 734 | - | 7,465 | 257 | - | -991 | - | 7,465 |
| $2003 / 2004$ | 800 | - | 11,536 | 610 | 14 | $-1,424$ | - | 11,536 |
| $2004 / 2005$ | 801 | 41 | 12,702 | - | - | -801 | - | 12,743 |
| $2005 / 2006$ | 821 | 150 | 9,494 | 799 | - | -1770 | - | 9,494 |
| $2006 / 2007$ | - | - | 6,944 | 518 | 5 | -523 | - | 6,944 |

Table 4.2.1.1. Celtic Sea \& Division VIIj herring. Comparison of age distributions (percentages) in the catches of Celtic Sea and VIIj herring over the time series.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1958 | 1 | 3 | 25 | 20 | 10 | 18 | 12 | 7 | 4 |
| 1959 | 1 | 27 | 2 | 20 | 12 | 6 | 19 | 4 | 8 |
| 1960 | 2 | 53 | 18 | 3 | 10 | 3 | 4 | 3 | 3 |
| 1961 | 3 | 22 | 44 | 8 | 3 | 7 | 4 | 2 | 7 |
| 1962 | 1 | 16 | 17 | 41 | 7 | 3 | 7 | 3 | 5 |
| 1963 | 0 | 52 | 13 | 4 | 21 | 3 | 1 | 3 | 3 |
| 1964 | 12 | 25 | 28 | 11 | 3 | 14 | 2 | 1 | 4 |
| 1965 | 0 | 56 | 8 | 13 | 3 | 4 | 10 | 1 | 6 |
| 1966 | 5 | 15 | 46 | 8 | 10 | 4 | 3 | 7 | 3 |
| 1967 | 5 | 26 | 13 | 32 | 6 | 6 | 3 | 4 | 4 |
| 1968 | 8 | 35 | 25 | 7 | 14 | 3 | 3 | 1 | 3 |
| 1969 | 4 | 40 | 24 | 14 | 5 | 8 | 2 | 1 | 1 |
| 1970 | 1 | 24 | 33 | 17 | 12 | 5 | 4 | 1 | 2 |
| 1971 | 8 | 15 | 24 | 27 | 12 | 7 | 3 | 3 | 1 |
| 1972 | 4 | 67 | 9 | 8 | 7 | 2 | 1 | 1 | 0 |
| 1973 | 16 | 26 | 38 | 5 | 7 | 4 | 2 | 2 | 1 |
| 1974 | 5 | 43 | 17 | 22 | 4 | 4 | 3 | 1 | 1 |
| 1975 | 18 | 22 | 25 | 11 | 13 | 5 | 2 | 2 | 2 |
| 1976 | 26 | 22 | 14 | 14 | 6 | 9 | 4 | 2 | 3 |
| 1977 | 20 | 31 | 22 | 13 | 4 | 5 | 3 | 1 | 1 |
| 1978 | 7 | 35 | 31 | 14 | 4 | 4 | 1 | 2 | 1 |
| 1979 | 21 | 26 | 23 | 16 | 5 | 2 | 2 | 1 | 1 |
| 1980 | 11 | 47 | 18 | 10 | 4 | 3 | 2 | 2 | 1 |
| 1981 | 40 | 22 | 22 | 6 | 5 | 4 | 1 | 0 | 1 |
| 1982 | 20 | 55 | 11 | 6 | 2 | 2 | 2 | 0 | 1 |
| 1983 | 9 | 68 | 18 | 2 | 1 | 0 | 0 | 1 | 0 |
| 1984 | 11 | 53 | 24 | 9 | 1 | 1 | 0 | 0 | 0 |
| 1985 | 14 | 44 | 28 | 12 | 2 | 0 | 0 | 0 | 0 |
| 1986 | 3 | 39 | 29 | 22 | 6 | 1 | 0 | 0 | 0 |
| 1987 | 4 | 42 | 27 | 15 | 9 | 2 | 1 | 0 | 0 |
| 1988 | 2 | 61 | 23 | 7 | 4 | 2 | 1 | 0 | 0 |
| 1989 | 5 | 27 | 44 | 13 | 5 | 2 | 2 | 0 | 0 |
| 1990 | 2 | 35 | 21 | 30 | 7 | 3 | 1 | 1 | 0 |
| 1991 | 1 | 40 | 24 | 11 | 18 | 3 | 2 | 1 | 0 |
| 1992 | 8 | 19 | 25 | 20 | 7 | 13 | 2 | 5 | 0 |
| 1993 | 1 | 72 | 7 | 8 | 3 | 2 | 5 | 1 | 0 |
| 1994 | 10 | 29 | 50 | 3 | 2 | 4 | 1 | 1 | 0 |
| 1995 | 6 | 49 | 14 | 23 | 2 | 2 | 2 | 1 | 1 |
| 1996 | 3 | 46 | 29 | 6 | 12 | 2 | 1 | 1 | 1 |
| 1997 | 3 | 26 | 37 | 22 | 6 | 4 | 1 | 1 | 0 |
| 1998 | 5 | 34 | 22 | 23 | 11 | 3 | 2 | 0 | 0 |
| 1999 | 11 | 27 | 28 | 11 | 12 | 7 | 1 | 2 | 0 |
| 2000 | 7 | 58 | 14 | 9 | 4 | 5 | 2 | 0 | 0 |
| 2001 | 12 | 49 | 28 | 5 | 3 | 1 | 1 | 0 | 0 |
| 2002 | 6 | 46 | 32 | 9 | 2 | 2 | 1 | 0 | 0 |
| 2003 | 3 | 41 | 27 | 16 | 6 | 4 | 3 | 0 | 1 |
| 2004 | 5 | 10 | 50 | 24 | 9 | 2 | 1 | 0 | 0 |
| 2005 | 19 | 38 | 7 | 23 | 9 | 2 | 1 | 0 | 0 |
| 2006 | 3 | 58 | 19 | 4 | 11 | 4 | 1 | 0 | 0 |

Table 4.2.1.2. Celtic Sea \& Division VIIj herring. Length frequency distributions of the Irish catches (raised numbers in '000s) in the 2006/2007 season in the Celtic Sea and VIIj fishery.


Table 4.2.3.1 Celtic Sea \& Division VIIj (2006/2007). Sampling intensity of commercial catches.

| ICES area | Year | Quarter | Landings (t) | No. Samples | Aged | . Measured | Aged/1000 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| VIIg | 2006 | 3 | 2957 | 13 | 893 | 3532 | 302 |
| VIIg | 2006 | 4 | 193 | 2 | 170 | 627 | 882 |
| VIIg | 2007 | 1 | 1624 | 9 | 658 | 2494 | 405 |
| Sub-total |  |  | 4773 | 24 | 1721 | 6653 | 1589 |
| VIIaS_inside harbour | 2006 | 4 | 568 | 30 | 1159 | 3328 | 2041 |
| VIIaS_outside harbour | 2006 | 4 | 1186 | 13 | 751 | 1897 | 633 |
| VIIaS_inside harbour | 2007 | 1 | 224 | 4 | 250 | 693 | 1116 |
| Sub-total |  |  | 1978 | 47 | 2160 | 5918 | 3790 |
| VIIj | 2006 | 4 | 135 | 2 | 152 | 398 | 1122 |
| Total Celtic Sea |  |  | 6887 | 73 | 4033 | 12969 | 6501 |

Table 4.2.3.2 Celtic Sea \& Division VIIj herring. Rudimentary history of the Irish fishery since 1958.

| Time period | 1958-1977 | 1977-1983 | 1983-1997 | 1998-2004 | 2004-2007 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Type of fishery | Cured fish | Closure | Herring roe | Fillet/whole fish Fillet/whole fish |  |
| Quality of catch data | High | Medium | Low | Medium/low | High |
| Source of catch data | Auction data | Auction data | Skipper estimate | Skipper estimate | Weighbridge landings |
| Discard Levels | Low | Low | High | Medium | Medium |
| Incentive to discard | None | None | Maturity stage | Size grade, market vs. quota |  |
| Alloowance for water* | na | na | na | 20\%* | 2\%* |

* RSW only. These vessels are more dominant in recent years.

Table 4.3.1.1. Celtic Sea \& Division VIIj herring. Acoustic surveys of Celtic Sea and VIIj herring, by season. Number of surveys per season and type indicated along with biomass and SSB estimates. Shaded sections show surveys not used in tuning.
$\left.\begin{array}{ll|cc}\hline & & & \text { Old } \\ \text { Season } & \text { No. Type } & \text { Revised } \\ \text { SSB }\end{array}\right]$

Table 4.3.1.2. Celtic Sea \& Division VIIj herring. Revised acoustic index of abundance. Total stock numbers-at-age ( $10^{6}$ ) estimated using combined acoustic surveys (age refers in winter rings, biomass and SSB in 000's tonnes).

|  | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 0 | 202 | 3 | - | 0 | - | 25 | 40 | 0 | 24 | - | 2 | - |
| 1 | 25 | 164 | - | 30 | - | 102 | 28 | 42 | 13 | - | 65 | 21 |
| 2 | 157 | 795 | - | 186 | - | 112 | 187 | 185 | 62 | - | 137 | 211 |
| 3 | 38 | 262 | - | 133 | - | 13 | 213 | 151 | 60 | - | 28 | 48 |
| 4 | 34 | 53 | - | 165 | - | 2 | 42 | 30 | 17 | - | 54 | 14 |
| 5 | 5 | 43 | - | 87 | - | 1 | 47 | 7 | 5 | - | 22 | 11 |
| 6 | 3 | 1 | - | 25 | - | 0 | 33 | 7 | 1 | - | 5 | 1 |
| 7 | 1 | 15 | - | 24 | - | 0 | 24 | 3 | 0 | - | 1 | - |
| 8 | 2 | 0 | - | 4 | - | 0 | 15 | 0 | 0 | - | 0 | - |
| 9 | 2 | 2 | - | 2 | - | 0 | 52 | 0 | 0 | - | 0 | - |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Abundance | 469 | 1338 | - | 656 |  | 256 | 681 | 423 | 183 |  | 312 | 305 |
| SSB | 36 | 151 |  | 100 |  | 20 | 95 | 41 | 20 |  | 33 | 36 |
| CV | 53 | 26 |  | 36 |  | 100 | 88 | 49 | 34 | 48 | 35 |  |

Table 4.6.1.1. Celtic Sea and VIIj herring.

| Cohort | Z (2-7 ring) | Cohort | Z (2-7 ring) |
| :--- | :---: | :---: | :---: |
|  |  |  |  |
| 1956 | 0.39 | 1977 | 1.09 |
| 1957 | 0.37 | 1978 | 0.84 |
| 1958 | 0.31 | 1979 | 0.93 |
| 1959 | 0.42 | 1980 | 0.75 |
| 1960 | 0.22 | 1981 | 0.75 |
| 1961 | 0.47 | 1982 | 0.65 |
| 1962 | 0.30 | 1983 | 0.63 |
| 1963 | 0.50 | 1984 | 0.50 |
| 1964 | 0.62 | 1985 | 0.66 |
| 1965 | 0.71 | 1986 | 0.62 |
| 1966 | 0.66 | 1987 | 0.76 |
| 1967 | 0.51 | 1988 | 0.58 |
| 1968 | 0.93 | 1989 | 0.73 |
| 1969 | 0.82 | 1990 | 0.57 |
| 1970 | 0.76 | 1991 | 0.65 |
| 1971 | 0.55 | 1992 | 0.77 |
| 1972 | 0.51 | 1993 | 0.90 |
| 1973 | 0.43 | 1994 | 0.73 |
| 1974 | 0.68 | 1995 | 0.80 |
| 1975 | 0.86 | 1996 | 1.02 |
| 1976 | 1.12 | 1997 | 0.88 |

Table 4.6.2.1. . Celtic Sea and VIIj herring. Settings used in exploratory VPA type assessments.

| Name | ICA Base case | ICA 7+ | ICA Sep=8 | ICA S=0.7 | XSA low |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |
| Separable period | 6 | 6 | 8 | 6 | - |
| Selection at oldest true age | 1 | 1 | 1 | 0.7 | - |
| Plus group | 9 ring | 7 ring | 9 ring | 9 ring | 9 ring |
| Mean F age range | $2-7$ | $2-5$ | $2-7$ | $2-7$ | $2-7$ |
| Shrinkage | No | No | No | No | Low (1.5) |
|  |  |  |  |  |  |

Table 4.6.2.2. Celtic Sea and VIIj herring. Input data for CSA trial assessment. CatRec and CatFul refer to catch numbers at 2 -ring and $3+$ ring respectively. Urec and Ufull refer to survey abundance at 2 -ring and $3+$ ring respectively. Wrec and Wfull refer to weights in the spawning stock for 2 -ring and $3+$ ring respectively.

| Year | CatRec |  | CatFull | Urec | Ufull |  | Wrec |  | Wfull |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| 1995 | 9450 | 71590 | 157.3 | 85.2 | 0.126 | 0.173008 |  |  |  |
| 1996 | 3476 | 68496 | 795.3 | 375.6 | 0.118 | 0.15978 |  |  |  |
| 1997 | 3849 | 101520 |  |  | 0.124 | 0.152527 |  |  |  |
| 1998 | 5818 | 75959 | 186.3 | 440.4 | 0.121 | 0.164057 |  |  |  |
| 1999 | 14274 | 79548 |  |  | 0.12 | 0.164147 |  |  |  |
| 2000 | 9953 | 46029 | 111.9 | 15.6 | 0.111 | 0.165149 |  |  |  |
| 2001 | 15724 | 49789 | 186.7 | 426.2 | 0.107 | 0.150619 |  |  |  |
| 2002 | 3495 | 27318 | 185.2 | 196.9 | 0.115 | 0.150171 |  |  |  |
| 2003 | 2711 | 51554 | 61.7 | 83 | 0.1 | 0.137398 |  |  |  |
| 2004 | 4276 | 78895 |  |  | 0.127 | 0.14077 |  |  |  |
| 2005 | 15419 | 34276 | 137.1 | 110 | 0.103 | 0.147259 |  |  |  |
| 2006 | 1460 | 22642 | 210.5 | 73.3 | 0.104 | 0.136832 |  |  |  |

Table 4.6.2.3. Celtic Sea and VIIj herring. WINBUGS code used for exploration of Celtic Sea herring

```
model
for (i in 1:I3) {
FAV[i] ~ dunif(.001,2) # i1 number of years of catch
# selection function priors - alternative to current selection
S1C~dunif(0.1,6) # catch ojive 50% age
S2C~dunif(0.2,6) # catch ojive 95% age - S1C
# Define the priors for survey Q values coefficients of
proportionality
for (j in 1:ACA) {
QAC[j]~dunif(.0001,20) # fit 1.36,
for (i in 1:(I3-1)) {
Nstar[i] ~ dnorm(1000,.000000000000064)
for (i in 1:I2){
Nin[i]<-1000*pow(10,i/5)
Nvar[i]<-.00001/pow(Nin[i]/4,2)
Nstar2[i] ~ dnorm(Nin[i],Nvar[i])
# Define the observation priors
tauy ~ dgamma(0.001,0.001)
sigy <- pow(1/tauy,.5)
# separate variance for Acoustic survey if required
#tauAC ~ dgamma(0.001,0.001)
#sigAC <- pow(1/tauAC,.5)
# constrained flexibility in selection - use next line to be fully
flexible
taus <- 100 # vary to change flexibility of separable model (low
values increase flexibility)
#taus ~ dgamma(0.001,0.001)
sigs <- pow(1/taus,.5)
################# main algorithm
## Selection curve as 2 parameter logistic function with random walk -
constrained to limit flexibility
S1CV[1]<-S1C
S2CV[1]<-S2C
for (i in 2:I3) {
ch1[i]~dnorm(0,taus)
ch2[i]~dnorm(0,taus)
ch2s[i]<-max(0.05,ch2[i])
S1CV[i]<-S1CV[i-1]+ch1[i]
S2CV[i]<-S2CV[i-1]+ch2s[i]
for (i in 1:I3 ){
for (j in 1:I2) {
FA[j,i]<-1/(1+exp(-2.944439*(age[j]-S1CV[i])/(S2CV[i])))
# selection pattern for catch
FAP[i]<-1/(1+exp(-2.944439*(agep-S1CV[i])/(S2CV[i])))
######### population component of the likelihood
# Define the system process for the population data
# stop any negative population numbers
for (i in 2:I3) {
N[1,i]<-max(Nstar[i-1],10)
for (i in 1:I2) {
N[i,1]<-max(Nstar2[i],10)
### set op selection period first
# start with matrix of Fs
for (i in 1:I3){
for (j in 1:I2){
F[j,i]<-FA[j,i]*FAV[i] # fishing mortality
INTF[j,i]<-F[j,i]/FA[j,i]
FP[i]<-FAP[i]*FAV[i] # fishing mortality
#Calculate N for ages 2 and greater and years after first year
```

```
for (i in 2:I3){
for (j in 2:I2){
N[j,i]<-N[j-1,i-1]* exp(F[j,i]+M[j,i])
for (i in 1:I3){
NP[i]<-CANUMP[i]*(FP[i]+MP[i])/FP[i]/(1-exp(-FP[i]-MP[i]))
#Then VPA part start with Ns age 0 to max age minus 2
#Then get Fs from Ns
# Mean F to set F on oldest real age and plus group
for (i in (I3+1):I1){
for (j in 2:(I2)){
N[j,i]<-N[j-1,i-1]* exp(M[j,i])+CANUM[j,i]* exp(M[j,i]/2)
F[j,i]<-log(N[j,i]/N[j-1,i-1])-M[j,i]
INTF[j,i]<-F[j,i]/FA[j,I3]
# calculate mean F and use selection to get F oldest real age and plus
group
FAV[i]<-mean(INTF[2:(I2-1),i])
# set Fs
F[1,i]<-FAV[i]*FA[1,I3]
FP[i]<-FAV[i]*FAP[I3]
# then set Ns for oldest ages
N[1,i]<-CANUM[1,i]*(F[1,i]+M[1,i])/F[1,i]/(1-exp(-F[1,i]-M[1,i]))
NP[i]<-CANUMP[i]*(FP[i]+MP[i])/FP[i]/(1-exp(-FP[i]-MP[i]))
# now cycle back in years
## Observation / objective function
for (i in 1:I1){
for (j in 1:I2){
SSBa[j,i]<-N[j,i]*exp((-F[j,i]*FPROP[j,i]-
M[j,i]*MPROP))*WEST[j,i]*MATPROP[j,i] ## at spawning time
SSB[i]<-sum(SSBa[,i])+NP[i]*exp((-FP[i]*FPROPP[i]-
MP[i]*MPROP))*WESTP[i]*MATPROPP[i] ## at spawning time
Fbar[i]<-(F[4,i]+F[5,i]+F[6,i]+F[7,i])/4 #### hard wired 1-4 here
should be flexible
## Acoustic survey
for (i in 1:ACY) {
for (j in 1:ACA) {
ModAC[j,i]<- log(QAC[j]*N[ACAind[j],ACYind[i]]) # log N with
constant multiplier at correct time of year set to 0 - 1st jan
ObsAC[j,i]~dnorm(ModAC[j,i],tauy)
# 2 Catch ##### assuming 25 survey values !!!!
for (i in 1:I3){
for (j in 1:(I2-1)){
ObsCatchMod[j,i]<-log(N[j,i]*F[j,i]/(F[j,i]+M[j,i])*(1-exp(-F[j,i]-
M[j,i])))
ObsCatch[j,i] ~ dnorm(ObsCatchMod[j,i],tauy)
# End of model
```

Table 4.6.2.4. Celtic Sea and VIIj herring. Time periods in the history of the stock.

|  | $1958-1972$ | $1973-1978$ | $1978-1980$ | $1981-1983$ | $1983-1995$ | $1996-1997$ | $1998-2004$ | $2004-2007$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
| MW (2-ring) kg | 0.146 | 0.181 | 0.179 | 0.158 | 0.135 | 0.121 | 0.115 | 0.112 |
| ML 2-ring (cm) | $\sim 25.5$ | 27.3 | 27.2 | 26.85 | - | - | - | 23.5 |
| Z (cohort catch curve) | 0.22 to 0.93 | 0.42 to 1.12 | 0.74 to 0.93 | 0.62 to 0.74 | $0.49-$ to 0.89 | 88 | to1.01 | $\sim 0.48$ |
| GM recruitment $10^{6}$ | 493 | 180 | 168 | 587 | 534 | 484 | 314 | - |
| Recruitment anomaly | -1.3 to 2.3 | -1.2 to -0.7 | -0.4 to 0.8 | 0.2 to 1.6 | -1.2 to 2.4 | -0.8 to 0 | -1 to 0.7 | - |
| SSB (000 t) | 79 to 126 | 29 to 59 | $28-$ to 0 | 32 to 63 | 63 to 81 | 52 to 59 | 22 to 44 | ? |
| F | 0.16 to 0.64 | 0.34 to 0.61 | 0.34 to 0.65 | 0.59 to 0.8 | 0.36 to 0.95 | 0.46 to 0.58 | 0.4 to 1.2 | ? |



Figure 4.1.1.1a. Celtic Sea and VIIj herring, areas mentioned in the text and spawning boxes $A$, $B$ and $C$, south of Ireland. One of these boxes is closed each season, under EU legislation. 1 Courtmacsherry, 2 Cork Harbour, 3 Daunt Rock, 4 Kinsale Gas Field (Rigs), 5 Labadie Bank, 6 Kinsale, 8 Waterford Harbour, 9, Baginbun Bay, 10, Tramore Bay/ Dunmore East, 11, Ballycotton Bay, 12, Valentia Island, 13 Kerry Head to Loop Head, 14, The Smalls. The spawning boxes A-C correspond to ICES Divisions VIIj, VIIg and VIIaS respectively.


Figure 4.1.1.1b. Celtic Sea and VIIj herring, Location of non-spawning (open symbol) and spawning (closed symbol) herring in the Celtic Sea and SW of Ireland. Based on expert fishermens' personal information.


Figure 4.1.2.1 Celtic Sea and Division VIIj - working group estimates of herring landings per season.


Figure 4.2.1.1. Celtic Sea and Division VIIj. Catch numbers at age standardised by yearly mean.


Figure 4.2.1.2. Celtic Sea and Division VIIj - percentage age composition by metier (ICES Division and quarter).


Figure 4.1.3.2. Celtic Sea and VIIj herring. Irish official herring catches by statistical rectangle in 2006/2007.


Figure 4.3.1.1a Celtic Sea and VIIj herring. Celtic Sea and Division VIIj acoustic survey 2006, survey track and haul positions from acoustic survey, October 2006.


Figure 4.3.1.1b. Celtic Sea and VIIj herring. Celtic Sea and Division VIIj acoustic survey 2006, total Sa values attributed to herring.


Figure 4.3.1.2. Celtic Sea and VIIj herring. The percentage age composition in the survey and the commercial fishery 2006.


Figure 4.4.1. Celtic Sea and VIIj herring, trends over time in mean weights in the catch.


Figure 4.4.2. Celtic Sea and VIIj herring, trends over time in mean weights in the stock at spawning time.


Figure 4.6.1.1. Celtic Sea and VIIj herring. Log catch ratios (above) and log catch ratios smoothed with a 4 year moving average for each age group for the time series 1958-2006.


Figure 4.6.1.2. Celtic Sea and VIIj herring. Cohort catch curves for the time series of catch at age data. Age in winter rings on the horizontal axis and log transformed catch numbers at age on the vertical axis.


Figure 4.6.1.3. Celtic Sea and VIIj herring. Total mortality ( $Z$ ) estimated from cohort catch curves (2-7 ringer) for cohorts 1958 to 1997.


Figure 4.6.1.4. Celtic Sea and VIIj herring. Cohort catch curves (2-5 ringer), averaged over several year classes, from catch at age data.


Figure 4.6.1.5. Celtic Sea and VIIj herring. Cohort catch curves (2-5 ring) based on acoustic survey abundance. Upper panel shows means for two periods, and below for three time periods, over the same series of surveys.


Figure 4.6.2.1. Celtic Sea and VIIj herring. Separable model residuals for four ICA exploratory assessments.


Figure 4.6.2.2. Celtic Sea and VIIj herring. Survey index residuals from four ICA exploratory runs.


Figure 4.6.2.3. Celtic Sea and VIIj herring. Log catchability residuals for XSA run, with low shrinkage.


Figure 4.6.2.4. Celtic Sea and VIIj herring. Exploratory assessment using ICA and XSA.



Figure 4.6.2.5. Celtic Sea and VIIj herring. Results of CSA trial assessment. Top panel, total biomass (2+), compared with ICA Base Case (spaly). Middle panel recruits (2-ring) estimated by acoustics and CSA estimates (solid line). Bottom panel, 3+ ruing acoustic index and estimated by acoustics and CSA estimates (solid line.


Figure 4.6.2.6. Celtic Sea and VIIj herring. WINBUGS estimates of a) SSB and b) mean F ages 2-5 showing the large uncertainty in SSB in the final year and recent decline in $F$, c) heavily constrained flexible selection (each bar is an age (1-8) by a year(1995 to 2006) sequentially by year in groups of ages, pattern rising with age from 1 to 8 , with declining means and wider intervals in later years at age $\mathbf{1}$; higher values and narrower intervals at age $\mathbf{8}$ in later years.


Figure 4.6.2.7. Celtic Sea and VIIj herring. Comparison of model fit a) log modelled catch on observed catch, b) log modelled survey estimate on log survey observation. The variance of the survey is greater than variance of the catch.




Figure 4.6.2.8. Celtic Sea and VIIj herring. Uncertainty around recruitment, SSB and mean F (2-7 ring) from ICA base case trial assessment. F transformed by addition of 1 and $\log$ transformed.



Figure 4.6.2.9. Celtic Sea and VIIj herring. Residual patterns around the stock recruit model fit. Upper panel, Beverton and Holt model and R/SSB residuals. Lower panel, Ricker and Hockey Stick Model.


| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(2-7) | 1.0000 | 0.4196 |
| FMax | $>=1000000$ | 0.0000 |
| F0.1 | 0.4454 | 0.1869 |
| F35\%SPR | 0.4578 | 0.1921 |
| Flow | 0.2054 | 0.0862 |
| Fmed | 0.6631 | 0.2783 |
| Fhigh | 3.0566 | 1.2827 |

Figure 4.8.1. Celtic Sea and VIIj herring. Yield per recruit and spawners per recruit analysis.




Figure 4.10.1. Celtic Sea and VIIj herring. Retrospective pattern in F, SSB and recruitment from ICA base case run.

## 5 West of Scotland Herring

### 5.1 The Fishery

### 5.1.1 ACFM Advice Applicable to 2006 and 2007

ACFM reported in 2006 that the state of the stock was uncertain. Exploratory assessments confirmed earlier perceptions of a lightly exploited stock ( $\mathrm{F}<=0.2$ ), but the level of the current biomass was uncertain. The recent level of fishing mortality was felt to be low and decreasing. The SSB, although uncertain was around $\mathbf{B}_{\mathrm{pa}}$. Given that the perception of the stock was the same as the previous year, the 2006 TAC should also be applicable in 2007.

There are no explicit management objectives for this stock. A $\mathbf{B}_{\mathrm{lim}}$ of 50000 t has been agreed by ACFM for this stock. A candidate HCR (see below) was presented by ACFM in 2005 with the statement that it "seems to maintain the stock inside precautionary limits" and ACFM agreed that it might be adopted subject to an evaluation of a year-on-year TAC constraint.

| $\mathrm{F}=0.25$ | if $\mathrm{SSB}>75000 \mathrm{t}$ | Optional year on year TAC constraint. |
| :--- | :--- | :--- |
| $\mathrm{F}=0.2$ | if $\mathrm{SSB}<75000 \mathrm{t}$ | No constraint on TAC. |
| $\mathrm{F}=0$ | if SSB falls below $\mathrm{B}_{\mathrm{lim}}$. |  |

The agreed TAC for 2007 is 34000 t , which is in accordance with the HCR above. The TAC in 2006 was 34000 t .

### 5.1.2 The VIa (North) Fishery

Historically, catches have been taken from this area by three fisheries.
i) A Scottish domestic pair trawl fleet and the Northern Irish fleet operated in shallower, coastal areas, principally fishing in the Minches and around the Island of Barra (Figure 5.1) in the south; younger herring are found in these areas. This fleet has reduced in recent years.
ii ) The Scottish single boat trawl and purse seine fleets, with refrigerated seawater tanks, targeting herring mostly in the northern North Sea, but also operated in the northern part of VIa (N). This fleet now operates mostly with trawls but many vessels can deploy either gear.
iii ) An international freezer-trawler fishery has historically operated in deeper water near the shelf edge where older fish are distributed. These vessels are mostly registered in the Netherlands, Germany, France and England but most are Dutch owned.

In recent years the catch of these last two fleets has become more similar. In 2006 the dominant year classes were 1999 and 2000 ( 6 and 5 ringers respectively). It appears that the 2001 year class is not strong as was originally supposed, but relatively weak.

In 2006, the Scottish trawl fleet fished in areas similar to the freezer trawler fishery, and not in the coastal areas in the southern part of VIa (N). The Northern Irish fleet fished in both the north and the south of VIa (N). In contrast to most of the previous years' fisheries, in 2006 $98 \%$ of the fishery was prosecuted in quarter 3 and was distributed only in the northern part of the area. In the past there has been a much more even distribution of effort, both temporally and spatially.

As a result of perceived problems of area misreporting of catch from IVa into VIa (N), Scotland introduced a fishery regulation in 1997 with the aim to improve reporting accuracy. Under this regulation, Scottish vessels fishing for herring were required to hold a license either
to fish in the North Sea or in the west of Scotland area (VIa (N)). Only one licensed option could be held at any one time. However in 2004, the requirement to carry only a single licence was rescinded. Area misreporting of catch taken in area IVa into area VIa (N) then increased in 2004 and continued in 2005. It is possible, therefore, that the relaxation of this single area licence contributed to a resurgence in area misreporting. In 2006, however, there was no misreporting from IVa into VIa (N). New sources of information on catch misreporting from the UK became available in 2006 (changes to catches will be discussed in the section below). This information was associated with a stricter enforcement regime that may be responsible for the lack of that area misreporting in 2006.

### 5.1.3 Catches in 2006 and Allocation of Catches to Area for VIa (North)

For 2006, the preliminary report of official catches corresponding to the VIa (N) herring stock unit total 34230 t , compared with the TAC of 34000 t . The Working Group's estimates of area misreported and unallocated catches are 6884 t . An additional 163 t of herring has been reported as discarded. At such a low level currently, discarding is not perceived to be a problem.

The Working Group’s best estimate of removals from the stock in 2006 is 27346 t (Table 5.1.1).

There were three revisions to the catch data time series in 2006. An incorrect allocation of fish to the plus group in the Dutch catches in 2004 and 2005 affected the 2004 VIaN catch and the allocations for that year had to be recalculated. Landings data were also revised with respect to reallocation of catches from area VIaN to VIaS, for the years 2000-2005 (see Section 6.1.2). Thirdly, a readjustment of catch figures was necessary from 2001 to 2004 in light of new information on misreporting from the UK. The resulting changes to the catch figures are documented in Table 5.1.1.

### 5.2 Biological composition of the catch

Catch and sample data, by country and by quarter, are detailed in Table 5.2.1. The number of samples used to allocate an age-distribution for the VIa (N) catches increased back up to the 2003 levels in 2006, after a few years of very poor sampling. There were 39 samples available in 2006, obtained from the Scottish, German and Dutch fleets. These were used to allocate a mean age-structure (weighted by the sampled catch) to unsampled catches, in the same quarter, or in adjacent quarters if no samples were available in the corresponding quarter. If no sampling data were available for a quarter, a mean age-structure of all samples from adjacent quarters was used. The allocation of age structures to unsampled catches, and the calculation of total international catch-at-age and mean weight-at-age in the catches were made using the 'sallocl' programme (Patterson, 1998) and compared to the new ICES Intercatch (Section 1.5.5). The samples obtained came from the major fisheries by fleet, area and season and are thought to be representative of the catches.

Catch in number-at-age information is given in Table 5.2.2. Three reasonable year classes can be seen clearly in the catch-at-age table: 1998, 1999 and 2000 at 7 -, 6- and 5-ring respectively in 2006. The 2001 year class, previously thought to be abundant is not at all dominant in the catch numbers-at-age in 2006 as 4 -ring fish. 1-ring herring in the catch are variable and are rarely representative of year class strength and are down-weighted in the assessment, see Section 5.6.

### 5.3 Fishery Independent Information

### 5.3.1 Acoustic Survey

The 2006 acoustic survey was carried out from 1- 21 July using a chartered commercial fishing vessel (MFV Enterprise). The total biomass estimate obtained, 471700 t , represents an almost doubling on the previous year (187500 in 2005) and is a return to the higher levels seen from 1998 to 2004 (Table 5.3.1). The abundance by year class is consistent with previous years and also with results from the adjacent North Sea area. The 1998 year class is depleted but still apparent; there were a significant number of 2 and 5 -ring fish seen on the survey (the 2003 and 2000 year classes respectively). The observed spatial distribution was different to previous years. Herring were found in some areas similar to those in previous surveys (Figures 2.3.1.5 and 5.1) i.e., to the west of the Hebrides, but there were two other areas of concentration in 2006 that are not usually seen, to the north of NW Scotland, around $59^{\circ} 30^{\prime} \mathrm{N}$, $5^{\circ} 30^{\prime} \mathrm{W}$ and in the south Minch to the north of the island of Coll. Further details are available in the Report of the Planning Group for Herring Surveys (ICES 2007/LRC:01). The same year classes seen in the catch can be seen clearly at 3-ring and older in the acoustic survey table. However, the 2003 year class at 2-ring in 2006 is considerably more dominant in the survey than in the catch and is the largest 2-ring abundance in the survey since the appearance of the abundant 2000 year class in 2003. To what extent this is a reflection of the high abundance of fish in the survey round Coll, a known area of abundance for younger herring, is unknown.

### 5.4 Mean weights-at-age and maturity-at-age

### 5.4.1 Mean Weight-at-age

Weights-at-age in the catches and weights-at-age in the stock from acoustic surveys are given in Table 5.4.1 and are used in the assessment. The weights-at-age in the catch are comparable to previous years. The weights-at-age in the stock, for 3 to 8-ringers, are higher than in 2005 and are consistent with the longer time series.

### 5.4.2 Maturity Ogive

The maturity ogive is obtained from the acoustic survey and collated in Table 5.4.2 for the period 1992 to 2006.

In 2006, maturity for 2 and 3-ring herring is more similar to the 2004 values than to the 2005 values with all herring above 3 -ring being mature.

### 5.5 Recruitment

There are no specific recruitment indices for this stock. Although both catch and acoustic survey have catches at 1-ring both the fishery and survey encounter this age group only incidentally. The first reliable appearance of a cohort appears at 2-ring in both the catch and the stock. Thus in predictions, estimates of both 1- and 2-ring herring numbers from the assessment need to be replaced for prediction years.

### 5.6 Assessment of VIa (North) herring

### 5.6.1 Data Exploration and Preliminary Modelling

In the 2007 HAWG, the VIa (North) assessment is a scheduled update assessment and there is no evidence that there are any specific modelling issues to be addressed. However, after worries about the low catch and survey values last year, the exclusion of the low survey values in the assessment was explored this year. This model has been explored in much detail in
recent years and is perceived to be reasonably well behaved with the settings used (see HAWG 2005). Therefore the model and the model settings used below are the same as last year's assessment and these will not be explored in detail this year. All exploratory assessments of the stock were carried out by fitting an integrated catch-at-age model (ICA version 1.4 w described in the methods section in the 2003 Working Group report (ICES 2003/ACFM:17, Section 1.6.1). An age-structured index was available from the acoustic survey from 1987, 1991-2006 (Section 5.3.1).

In 2006 there were several revisions to the historical data that required exploration (Tables 5.1.1, 5.2.2 and 5.4.1). An exploratory assessment was performed to investigate the effects these revisions had on our perception of the stock. The text table below shows the change in spawning stock biomass (SSB), F and total stock biomass (TSB) in the terminal year for last year's (the 2006 working group) assessment and the assessment repeated but with the revised input data. It also shows the overall change for the 2005 estimates from this year's assessment.

|  | SSB | F | TSB |
| :--- | ---: | ---: | ---: |
| 2006 assessment, using original data for 2005 estimates | 64110 | 0.203 | 94611 |
| 2006 assessment, using revised data for 2005 estimates | 71291 | 0.153 | 102437 |
| 2007 assessment, using revised data for 2005 estimates | 88261 | 0.126 | 111054 |

The catch revision resulted in a slightly lower catch over the period 2000 - 2005 and this gives a small upward revision (10\%) in SSB and consequent decrease in F for the assessment with revised data. In the 2007 assessment the increased biomass in the 2006 survey results in upward revision of the SSB value for 2005 (by 27\%) and a consequent decrease in F.

It was decided this year to explore the use of the two low survey values in the time series, from 1997 and 2005. It was decided to consider the inclusion or exclusion of both, as they are similar, rather than one or the other individually. The 1997 value (Table 5.3.1) has never been included in the tuning index. It is the lowest in the time series and was conducted almost a month earlier than all the other surveys. The 2005 survey value was the second lowest in the time series and coupled with a low level of catch. In last year's working group, exploration of the assessment showed quite different stock perceptions depending on whether the survey tuning index was included or excluded in the assessment runs. Both the 1997 and 2005 survey SSB estimates show the same relative year effect (Table 5.3.1).

With the revised data set ICA was then run for the time-series 1958-2006, to compare the exploratory model fits for this year. The full time series, 1957-2006, was unable to be used because, currently, ICA is only able to be run on a maximum time series of 49 years.

The separable model residual patterns for the two runs (excluding the 1997 and 2005 surveys and then including both surveys) are very similar (Figure 5.6.1). The magnitude and location of residuals shown in the bubble plots are consistent and the year residuals follow the same pattern. The age residuals values are all small and there are no trends with age. However, the values are slightly larger when both the 1997 and 2005 surveys are included.

The survey residuals patterns for the two runs are mostly similar (Figure 5.6.2). The magnitude and location of residuals shown in the bubble plots are mostly consistent and show strong year effects, whether the two surveys are included or excluded. In both runs it can be seen that the survey residuals show a better pattern in the period prior to 1998, with fewer year effects and less pattern in the distribution of positive and negative residuals. In the later period, for both runs, there is a twisting of the pattern, with a switch from strong negative to positive residuals for the early ages and vice versa for the older ages. This pattern is reflected in the year residuals and most likely caused by conflicting signals seen in the catch and survey
data (and explored in details in the 2006 HAWG - see the log catch-ratio plots in Figure 5.6.4. therein). Examination of the year residuals shows that the two low survey values produce residuals of the same magnitude to others in the time series. There is little difference between the two runs in terms of the residual plots.

A plot to compare the reference F (from the parameter estimates) in the terminal year (Figure 5.6.3) shows small differences when the two surveys are either excluded or included in the assessment. The run excluding the two surveys has a marginally wider confidence interval, and although the value of F is lower excluding the two low surveys the two values are essentially the same. The inclusion or exclusion also has a minimal effect on the estimate of q.

Figure 5.6.4 shows the values for SSB and F produced by the two assessment runs. There is a minimal difference between the two values for both SSB and F, with a higher SSB, and therefore lower F , with the run excluding the two low surveys. These differences are within the bounds of the confidence intervals of the assessment.

Retrospective analyses of the assessment from 2006 to 2002 were carried out, and are compared with the two runs excluding and including the 1997 and 2005 acoustic surveys. Figure 5.6 .5 shows the SSB, mean $F_{3-6}$ and recruitment from ICA assessments, with an 8 year separable period. Generally, in the year of assessment recruitment is very poorly estimated. However, in these assessments there is broad agreement in the patterns of recruitment. The retrospective patterns of SSB converge around 2000. The exploratory assessment values are broadly central within the retrospective pattern, i.e., there is no discernable bias. This, coupled with the patterns in the residuals and reference $F$ discussed above, suggests that the assessment is not sensitive to the inclusion or exclusion of the two low survey values in 1997 and 2005. It was therefore decided to include both survey values in future assessments.

### 5.6.2 Stock Assessment

This is an update assessment using the same settings as in 2006, with the 8 year separable period moved forward one year from 1998-2005 to 1999-2006, using the complete survey time series.

Assessment of the stock was carried out by fitting an integrated catch-at-age model (ICA version 1.4 w ). The model settings are the same as in 2005 and 2006. The run log is shown in Table 5.6.1. The input data are given in Tables 5.6.2 to 5.6.8. The output data are given in Tables 5.6.9 to 5.6.18.

The assessment results in an SSB for 2006 of $77800 t$ and a mean fishing mortality (3 to 6ringers) of 0.276 . Figure 5.6 .6 illustrates the stock trends from the assessment ( 8 year separable period). The model diagnostics (Tables 5.6.13 to 5.6.18 and Figure 5.6.7) show that the total residuals by age and year between the catch and separable model are reasonably trend-free. The acoustic survey residuals are of a higher magnitude than the catch residuals and show more evidence of year effects. There is also a switch from a strongly positive pattern in 2000 to some large negative values in 2005, discussed above in Section 5.6.1. The large 1998 year class is still reasonably abundant in the catch and survey data in 2006. The 2000 year class is most prevalent in the catch data (5-ringers). This year's estimate of SSB for 2005 is 88261 t , compared with 64110 t in last year's assessment run including the 2005 (but excluding the 1997) survey. The assessment run shows an increased catch, decreased recruitment and decreased SSB with an increased $F$ in the last four years.

The outcome of the assessment this year confirms earlier perceptions of a fairly lightly exploited stock although F is higher than in previous years ( $\mathrm{F}<=0.28$ ). This year the assessment of the current biomass is more certain than last year. Catch has increased on last year (almost doubled). The SSB has decreased by $45 \%$ since its previous high value in 2002, and by around $30 \%$ since 2004, likely a result of lower recruitment and an increased catch.

Recruitment for the 2001 year classes onwards shows the longest series of low recruitments in the time series (Table 5.6.14).

### 5.7 Short term projections

### 5.7.1 Deterministic short-term projections

Short-term projections were carried out using MFDP. Input data are stock numbers on $1^{\text {st }}$ January in 2007 from the 2006 ICA assessment (Section 5.6.2, Table 5.6.10), with geometric mean replacing recruitment for both 1- and 2-ring in 2007. In 2007, the geometric mean was calculated for the time series 1989 - 2005, a period showing lower productivity than earlier in time (see Section 1.8.3). This was felt necessary given that there is evidence for poor recruitment in recent year classes (Table 5.6.10). The selection pattern used is as estimated by ICA (Table 5.6.13). For the projections, data for maturity, natural mortality, mean weights-atage in the catch and in the stock are means of the three previous years (i.e., 2004-2006) (Table 5.7.1.1). Two scenarios for deterministic short-term projections are presented: F status quo and a second option with TAC constraint. The results of short-term projection is shown in the text table below, illustrating that at status quo F catches can be expected to be stable at around 21000 t .

| Scenario | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: |
| 1 - status quo F | $\mathrm{F}_{2007}=\mathrm{F}_{2006}=0.276$ <br> Status quo F $\text { Catch }=21832 \mathrm{t}$ | $\mathrm{F}_{2008}=\mathrm{F}_{2006}=0.276$ <br> Status quo F $\text { Catch }=20772 \text { t }$ | $\mathrm{F}_{2009}=\mathrm{F}_{2006}=0.276$ <br> Status quo F $\begin{aligned} & \text { Catch }=21044 \mathrm{t} \\ & \mathrm{SSB}=77309 \mathrm{t} \end{aligned}$ |
| TAC Constraint | $\mathrm{F}_{2007}=0.46$ <br> TAC Constraint <br> Catch $=34000 \mathrm{t}$ | $\mathrm{F}_{2008}=\mathrm{F}_{\mathrm{int}}=0.20$ <br> F management plan intermediate $\text { SSB = } 69015$ | $\mathrm{F}_{2009}=\mathrm{F}_{\text {int }}=0.20$ <br> F management plan intermediate $\mathrm{SSB}=75403 \mathrm{t}$ |

There is a proposed management plan (Section 5.1.1) that may be implemented in 2008. This relates to work carried out in the last 5 years by HAWG. It has therefore been included in the projections.

The results of the F status quo short-term projections can be seen in Tables 5.7.1.2 - 5.7.1.3. Table 5.7.1.2 shows single option predictions for 2008 and 2009. Table 5.7.1.3 shows the multiple options for 2008. SSB rises from approximately 74000 t in 2007 to around 77000 t in 2009. The results of the TAC constraint for 2007 short-term projections can be seen in Tables 5.7.1.4-5.7.1.5. With the current TAC of 34000 t , and the current decreased SSB in the assessment, F in 2007 is high, at $\mathrm{F}=0.46$, far higher than the proposed upper F value in the management plan of $F=0.25$. If the stock is subsequently fished even at $F=0.20$ in 2008, SSB will decrease to around 69000 t . Continued fishing at $\mathrm{F}=0.20$ in 2009 will lead to an increase to 75400 t . These projections may be considered to be conservative, based on a reduced geometric mean recruitment of 605 million. Table 5.6.10 allows a calculation of a value for 3 recent (2001 - 2003) year classes of around 265 million. However, the 2003 year class is well represented in the 2006 acoustic survey so the strength of the recruitment signal is unclear.

So, under both scenarios the assessment gives a starting position of SSB being below 75000 t ( $\mathrm{B}_{\text {trig }}$ in the proposed management rule). At F status quo the projection shows the stock rebuilding to an SSB around 77000 t in 2009 with a corresponding catch of 21000 t . At the current TAC though F is very high and the stock can only rebuild to around $\mathrm{B}_{\text {trig }}$ in 2009 at $\mathrm{F}=0.20$.

### 5.7.2 Yield-per-recruit

A yield-per-recruit analysis was carried out using MFYPR to provide yield-per-recruit plots for the data produced in the assessment run described above (Figure 5.7.2.1) The values for $\mathbf{F}_{0.1}$ and $\mathbf{F}_{\text {med }}$ ( 0.17 and 0.27 respectively) are very similar to last year's values derived which were the same regardless of whether the 2005 acoustic survey was included or excluded. These reference points therefore appear to be stable. These may be compared with the current F (2006 assessment) of 0.276 .

### 5.8 Medium term projections and HCR performance

Medium term projections were used extensively at the 2005 HAWG to evaluate HCRs for this area. This work was developed further through 2006 and is now published (Simmonds and Keltz 2007). There is no evidence that the stock diagnostics have changed, so the proposed rule (Section 5.1.1) should be adequate to protect the stock. Currently medium term management implications are not affected by the recommendations of WESTHER (Section 1.3.1).

### 5.9 Precautionary and yield based reference points

The biomass limit point $\mathrm{B}_{\lim }$ is $50,000 \mathrm{t}$. There are no agreed precautionary reference points for this stock. The proposed management rule has a $B_{\text {trig }}$ at 75000 t .

### 5.10 Quality of the Assessment

The HAWG considers the assessment this year to be as reliable as usual. This assessment has rather high variability, but this has been taken into account within the proposed HCR (Section 5.1.1). The influence of catch revisions was explored and seen to make little difference to the assessment outcome. Similarly, the behaviour of the two low survey SSB estimates was explored in the assessment through their inclusion or exclusion. Again, the assessment outcomes were very similar and well within the bounds of the confidence intervals of the assessment.

Retrospective analyses of the assessment from 2006 to 2002 were carried out, and are compared with the two runs excluding and including the 1997 and 2005 acoustic surveys. Figure 5.6.5 shows the SSB, mean $\mathrm{F}_{3-6}$ and recruitment from ICA assessments, with an 8 year separable period. In these assessments there is broad agreement in the patterns of recruitment. The retrospective patterns of SSB converge around 2000. The exploratory assessment values are broadly central within the retrospective pattern, i.e., there is no discernable bias. The results suggested that the assessment is not sensitive to the inclusion or exclusion of the two low survey values in 1997 and 2005. It was therefore decided to include both survey values in future assessments. The retrospective pattern supports the perception of a noisy but fairly well balanced assessment that has been assumed for the HCR considerations (Simmonds and Keltz 2007)

### 5.11 Management Considerations

In 2006, the stock was more heavily exploited than it has been since 1999. This recent increased $F$ is associated with: increased enforcement, reducing area misreporting from area IVa (Section 5.1.2), a roll-over TAC advised for 2007 due to some uncertainty with the assessment. Values since 2001 show the longest series of low recruitments in the time series (the 2001 to 2003 year classes (Table 5.6.14)), although the 2003 value is uncertain as it is seen as high in the survey and low in the fishery. Recruitment at 1wr in 2006 and 2007 is uncertain. In 2007, if the TAC, set as a roll-over TAC, is taken, F is predicted to reach $\mathrm{F}=0.46$. $F$ in 2006 was above the $\mathrm{F}=0.25$ recommended as the upper fishing mortality in the proposed
management plan. The short-term projections, based on a lower geometric mean (Sections 1.8.3 and 5.7) reflecting the current lower productivity of this stock, suggest that SSB will only rebuild to around the proposed $\mathrm{B}_{\text {trig }}$ of 75000 t in 2009 if F remains at $\mathrm{F}=0.20$ in both 2008 and 2009. Considering the roll-over TAC for 2007 and the high F in 2006, a 15\% restriction on TAC change is not advisable for 2008.

Following the recommendations from WESTHER (Hatfield et al. WD 2007) HAWG considers that in the absence of any evaluated and coordinated management strategy for the herring to the west of the British Isles, the current separation of management units (VIa(N), $\mathrm{VIa}(\mathrm{S})$, Irish Sea and Celtic Sea) affords the best possible protection for local spawning stocks. However it does not afford protection to the fish of one stock distributed in another management area at feeding time.

Provided both the spawning fisheries (VIa(S), Irish Sea and Celtic Sea) and the fishery in the mixing area (predominantly $\operatorname{VIa}(\mathrm{N})$ ) are maintained at an F that would be sustainable for each component, this should afford protection for these units, in the short term. HAWG considers that further work is required on examining the issues surrounding surveys, assessment and management of each of the current three management units to the north of the area. This can be initiated, partly through a new study group or study contract. It will be a number of years before ICES can provide a fully operational integrated strategy for these units. In this context ICES recommends that the previously endorsed plans for $\mathrm{VIa}(\mathrm{N})$ should be continued, until or unless some alternative strategy is found to be more useful.

Table 5.1.1 Herring in VIa (N). Catch in tonnes by country, 1983-2006. These figures do not in all cases correspond to the official statistics and cannot be used for management purposes.

| Country | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Denmark |  | 96 |  |  |  |  |  |  |
| Faroes | 834 | 954 | 104 | 400 |  |  |  |  |
| France | 1313 |  | 20 | 18 | 136 | 44 | 1342 | 1287 |
| Germany | 6283 | 5564 | 5937 | 2188 | 1711 | 1860 | 4290 | 7096 |
| Ireland |  |  |  | 6000 | 6800 | 6740 | 8000 | 10000 |
| Netherlands | 20200 | 7729 | 5500 | 5160 | 5212 | 6131 | 5860 | 7693 |
| Norway | 7336 | 6669 | 4690 | 4799 | 4300 | 456 |  | 1607 |
| UK | 31616 | 37554 | 28065 | 25294 | 26810 | 26894 | 29874 | 38253 |
| Unallocated | -4059 | 16588 | -502 | 37840 | 18038 | 5229 | 2123 | 2397 |
| Discards |  |  |  |  |  |  |  |  |

*WG estimate for 1997 has been revised according to the Bayesian assessment (see text Section 5.1.3 of 2000 report). ${ }^{\text {S }}$ Revised at HAWG 2007.

Table 5.2.1 Herring in VIa (N). Catch and sampling effort by nations participating in the fishery in 2006.

| PERIOD : 1 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Sampled | Official | No. of | No. | No. | SOP |
|  | Catch | Catch | samples | measured | aged | \% |
| Germany | 33.00 | 33.00 | 3 | 574 | 295 | 95.01 |
| Germany discard | 5.00 | 5.00 | 1 | 23 | 23 | 97.78 |
| Ireland | 0.00 | 632.00 | 0 | 0 | 0 | 0.00 |
| N. Ireland | 0.00 | 12.00 | 0 | 0 | 0 | 0.00 |
| Netherlands | 0.00 | 350.00 | 0 | 0 | 0 | 0.00 |
| Scotland | 0.00 | 261.00 | 0 | 0 | $\bigcirc$ | 0.00 |
| Scotland discard | 158.00 | 158.00 | 1 | 35 | 0 | 99.47 |
| Period Total | 196.00 | 1451.00 | 5 | 632 | 318 | 98.68 |
| Sum of Offical Catches: |  | 1451.00 |  |  |  |  |
| Unallocated Catch: |  | -944.00 |  |  |  |  |
| Working Group Catch: |  | 507.00 |  |  |  |  |

PERIOD : 2

| Country | Sampled |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Catch | Official | No. of | No. | No. | SOP |
| England \& Wales | 0.00 | 18.00 | samples | 0 | measured | aged |

PERIOD : 3

| Country | Sampled Catch | Official Catch | No. of samples | No. measured | No. aged | $\begin{gathered} \text { SOP } \\ \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| England \& Wales | 0.00 | 2854.00 | 0 | 0 | 0 | 0.00 |
| Faroes | 0.00 | 570.00 | 0 | 0 | 0 | 0.00 |
| France | 0.00 | 701.00 | 0 | 0 | 0 | 0.00 |
| Germany | 3119.00 | 3119.00 | 5 | 1440 | 111 | 100.42 |
| $N$. Ireland | 0.00 | 757.00 | 0 | 0 | 0 | 0.00 |
| Netherlands | 4438.00 | 6314.00 | 4 | 404 | 100 | 99.73 |
| Scotland | 14382.00 | 14382.00 | 25 | 4329 | 1061 | 99.99 |
| Period Total | 21939.00 | 28697.00 | 34 | 6173 | 1272 | 100.00 |
| Sum of Offical Catches |  | 28697.00 |  |  |  |  |
| Unallocated | Catch: | -1876.00 |  |  |  |  |
| Working Group Catch: |  | 26821.00 |  |  |  |  |

PERIOD : 4


Table 5.2.2 Herring in VIa (N). Estimated catch numbers-at-age (thousands), 1976-2006. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Age | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 69053 | 34836 | 22525 | 247 | 2692 | 36740 | 13304 | 81923 | 2207 | 40794 | 33768 |
| 2 | 319604 | 47739 | 46284 | 142 | 279 | 77961 | 250010 | 77810 | 188778 | 68845 | 154963 |
| 3 | 101548 | 95834 | 20587 | 77 | 95 | 105600 | 72179 | 92743 | 49828 | 148399 | 86072 |
| 4 | 35502 | 22117 | 40692 | 19 | 51 | 61341 | 93544 | 29262 | 35001 | 17214 | 118860 |
| 5 | 25195 | 10083 | 6879 | 13 | 13 | 21473 | 58452 | 42535 | 14948 | 15211 | 18836 |
| 6 | 76289 | 12211 | 3833 | 8 | 9 | 12623 | 23580 | 27318 | 11366 | 6631 | 18000 |
| 7 | 10918 | 20992 | 2100 | 4 | 8 | 11583 | 11516 | 14709 | 9300 | 6907 | 2578 |
| 8 | 3914 | 2758 | 6278 | 1 | 1 | 1309 | 13814 | 8437 | 4427 | 3323 | 1427 |
| 9 | 12014 | 1486 | 1544 | 0 | 0 | 1326 | 4027 | 8484 | 1959 | 2189 | 1971 |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| 1 | 19463 | 1708 | 6216 | 14294 | 26396 | 5253 | 17719 | 1728 | 266 | 1952 | 1193 |
| 2 | 65954 | 119376 | 36763 | 40867 | 23013 | 24469 | 95288 | 36554 | 82176 | 37854 | 55810 |
| 3 | 45463 | 41735 | 109501 | 40779 | 25229 | 24922 | 18710 | 40193 | 30398 | 30899 | 34966 |
| 4 | 32025 | 28421 | 18923 | 74279 | 28212 | 23733 | 10978 | 6007 | 21272 | 9219 | 31657 |
| 5 | 50119 | 19761 | 18109 | 26520 | 37517 | 21817 | 13269 | 7433 | 5376 | 7508 | 23118 |
| 6 | 8429 | 28555 | 7589 | 13305 | 13533 | 33869 | 14801 | 8101 | 4205 | 2501 | 17500 |
| 7 | 7307 | 3252 | 15012 | 9878 | 7581 | 6351 | 19186 | 10515 | 8805 | 4700 | 10331 |
| 8 | 3508 | 2222 | 1622 | 21456 | 6892 | 4317 | 4711 | 12158 | 7971 | 8458 | 5213 |
| 9 | 5983 | 2360 | 3505 | 5522 | 4456 | 5511 | 3740 | 10206 | 9787 | 31108 | 9883 |
|  | 1998 | 1999 | $2000{ }^{\text {s }}$ | $2001{ }^{\text {\$ }}$ | $2002{ }^{\text {\$ }}$ | $2003{ }^{\text {s }}$ | $2004{ }^{\text {§ }}$ | 2005 | 2006 |  |  |
| 1 | 9092 | 7635 | 3569 | 143 | 992 | 56 | 0 | 183 | 132 |  |  |
| 2 | 74167 | 35252 | 18162 | 81030 | 38482 | 33332 | 6844 | 9633 | 6691 |  |  |
| 3 | 34571 | 93910 | 17264 | 14943 | 93975 | 46866 | 22223 | 23237 | 9186 |  |  |
| 4 | 31905 | 25078 | 40674 | 9306 | 9014 | 53767 | 27815 | 20602 | 13645 |  |  |
| 5 | 22872 | 13364 | 12264 | 24482 | 18114 | 7463 | 45782 | 10238 | 41068 |  |  |
| 6 | 14372 | 7529 | 7121 | 9281 | 28016 | 4345 | 3916 | 9783 | 27782 |  |  |
| 7 | 8641 | 3251 | 3083 | 6625 | 9040 | 12818 | 7642 | 1015 | 20973 |  |  |
| 8 | 2825 | 1257 | 1452 | 4611 | 1548 | 9188 | 8481 | 1195 | 3042 |  |  |
| 9 | 3327 | 1089 | 456 | 1001 | 1423 | 1408 | 4008 | 1431 | 5089 |  |  |
| \$ revised at HAWG 2007 |  |  |  |  |  |  |  |  |  |  |  |

Table 5.3.1 Herring in VIa (N). Estimates of abundance from Scottish acoustic surveys. Thousands of fish at age and spawning biomass (SSB, tonnes). N.B. In this table "age" refers to number of rings (winter rings in the otolith).

*Biomass of 2+ ringers in November.
\# The 1997 survey is not on the same basis as the other years, it was conducted in June (all other surveys were carried out in July).

Table 5.4.1 Herring in VIa (N). Mean weights-at-age (g). N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Weights in the catch |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | $2001{ }^{\text {S }}$ | $2002^{\$}$ | $2003{ }^{\text {\$ }}$ | $2004{ }^{\text {S }}$ | 2005 | 2006 |
| 1 | 82 | 79 | 84 | 91 | 89 | 83 | 105 | 81 | 89 | 97 | 76 | 83 | 49 | 107 | 60 | 0 | 108 | 91 |
| 2 | 142 | 129 | 118 | 122 | 128 | 142 | 142 | 134 | 136 | 138 | 130 | 137 | 140 | 146 | 145 | 154 | 133 | 158 |
| 3 | 145 | 173 | 160 | 172 | 158 | 167 | 180 | 178 | 177 | 159 | 158 | 164 | 163 | 163 | 160 | 173 | 163 | 168 |
| 4 | 191 | 182 | 203 | 194 | 197 | 190 | 191 | 210 | 205 | 182 | 175 | 183 | 183 | 173 | 169 | 195 | 185 | 193 |
| 5 | 190 | 209 | 211 | 216 | 206 | 195 | 198 | 230 | 222 | 199 | 191 | 201 | 192 | 160 | 186 | 216 | 211 | 208 |
| 6 | 213 | 224 | 229 | 224 | 228 | 201 | 213 | 233 | 223 | 218 | 210 | 215 | 196 | 179 | 200 | 220 | 226 | 225 |
| 7 | 216 | 228 | 236 | 236 | 223 | 244 | 207 | 262 | 219 | 227 | 225 | 239 | 205 | 187 | 194 | 199 | 234 | 244 |
| 8 | 204 | 237 | 261 | 251 | 262 | 234 | 227 | 247 | 238 | 212 | 223 | 281 | 225 | 245 | 186 | 190 | 256 | 262 |
| 9+ | 243 | 247 | 271 | 258 | 263 | 266 | 277 | 291 | 263 | 199 | 226 | 253 | 272 | 281 | 294 | 311 | 250 | 275 |

\$ Revised at HAWG 2007

| Weight in the stock from acoustic surveys |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Historical | 1994 | 1995 | 1996 | 1997* | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 90 | 52 | 45 | 45 | 57 | 65 | 54 | 62 | 62 | 62 | 64 | 54 | 75 | 75 |
| 2 | 164 | 150 | 144 | 140 | 150 | 138 | 137 | 141 | 132 | 153 | 138 | 136 | 130 | 135 |
| 3 | 208 | 192 | 191 | 180 | 189 | 177 | 166 | 173 | 170 | 177 | 176 | 157 | 154 | 166 |
| 4 | 233 | 220 | 202 | 209 | 209 | 193 | 188 | 183 | 190 | 198 | 190 | 180 | 167 | 185 |
| 5 | 246 | 221 | 225 | 219 | 225 | 214 | 203 | 194 | 198 | 212 | 204 | 189 | 180 | 192 |
| 6 | 252 | 233 | 226 | 222 | 233 | 226 | 219 | 204 | 212 | 215 | 213 | 202 | 191 | 204 |
| 7 | 258 | 241 | 247 | 229 | 248 | 234 | 225 | 211 | 220 | 225 | 217 | 213 | 213 | 211 |
| 8 | 269 | 270 | 260 | 242 | 266 | 225 | 235 | 222 | 236 | 243 | 223 | 214 | 203 | 224 |
| 9+ | 292 | 296 | 293 | 263 | 287 | 249 | 245 | 230 | 254 | 259 | 228 | 206 | 228 | 231 |

[^9]Table 5.4.2 Herring in VIa (N). Maturity ogive used in estimates of spawning stock biomass taken from acoustic surveys. Values measured in 1997 were measured in June whilst other values are measured in July. The mean value 92-96 is used in the assessment for the years 1976-1991 and 1997.

| Year \Age <br> (Winter ring) | 2 | 3 | $>3$ |
| :--- | :--- | :--- | :--- |
| Mean 92-96 | 0.57 | 0.96 | 1.00 |
| 1992 | 0.47 | 1.00 | 1.00 |
| 1993 | 0.93 | 0.96 | 1.00 |
| 1994 | 0.48 | 0.92 | 1.00 |
| 1995 | 0.19 | 0.98 | 1.00 |
| 1996 | 0.76 | 0.94 | 1.00 |
| 1997 | 0.41 | 0.88 | 1.00 |
| 1998 | 0.85 | 0.97 | 1.00 |
| 1999 | 0.57 | 0.98 | 1.00 |
| 2000 | 0.45 | 0.92 | 1.00 |
| 2001 | 0.93 | 0.99 | 1.00 |
| 2002 | 0.92 | 1.00 | 1.00 |
| 2003 | 0.76 | 1.00 | 1.00 |
| 2004 | 0.83 | 0.97 | 1.00 |
| 2005 | 0.84 | 1.00 | 1.00 |
| 2006 | 0.81 | 0.97 | 1.00 |

Table 5.6.1. Herring in VIa (N). ICA run log for the maximum-likelihood ICA calculation for the 8 year separable period. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

```
Integrated Catch at Age Analysis
                    Version 1.4 w
                    K.R.Patterson
Fisheries Research Services
            Marine Laboratory
                Aberdeen
    Enter the name of the index file -->index.txt
canum.txt
weca.txt
    Stock weights in 2007 used for the year 2006
west.txt
    Natural mortality in 2007 used for the year 2006
natmor.txt
    Maturity ogive in 2007 used for the year 2006
matprop.txt
    Name of age-structured index file (Enter if none) : -->fleet.txt
    Name of the SSB index file (Enter if none) -->
No indices of spawning biomass to be used.
    No of years for separable constraint ?--> 8
    Reference age for separable constraint ?--> 4
    Constant selection pattern model (Y/N) ?-->y
    S to be fixed on last age ?--> 1.000000000000000
    First age for calculation of reference F ?--> 3
    Last age for calculation of reference F ?--> 6
    Use default weighting (Y/N) ?-->n
Enter relative weights at age
    Weight for age 1--> 0.100000000000000
    Weight for age 2--> 1.000000000000000
    Weight for age 3--> 1.000000000000000
    Weight for age 4--> 1.000000000000000
    Weight for age 5--> 1.000000000000000
    Weight for age 6--> 1.000000000000000
    Weight for age 7--> 1.000000000000000
    Weight for age 8--> 1.000000000000000
    Weight for age 9--> 1.000000000000000
Enter relative weights by year
    Weight for year 1999--> 1.000000000000000
    Weight for year 2000--> 1.000000000000000
    Weight for year 2001--> 1.000000000000000
    Weight for year 2002--> 1.000000000000000
    Weight for year 2003--> 1.000000000000000
    Weight for year 2004--> 1.000000000000000
    Weight for year 2005--> 1.000000000000000
    Weight for year 2006--> 1.000000000000000
Enter new weights for specified years and ages if needed
    Enter year, age, new weight or -1,-1,-1 to end. -1 -1 -1.000000000000000
    Is the last age of FLT01:West Scotland Summer Acoustic Surv a plus-group (Y--
>y
You must choose a catchability model for each index.
Models: A Absolute: Index = Abundance . e
    L Linear: Index = Q. Abundance . e
    P Power: Index = Q. Abundance^ K .e
```

    where Q and K are parameters to be estimated, and
    e is a lognormally-distributed error.
    Model for FLT01:West Scotland Summer Acoustic Surv is to be A/L/P ?-->L
There are 1 missing observations for fitting the separable model.
Fit a stock-recruit relationship (Y/N) ?-->n
Enter lowest feasible F--> 2.0000000000000000E-02
Enter highest feasible F--> 0.500000000000000

Table 5.6.1. continued.
Mapping the F-dimension of the SSQ surface
F SSQ

| +------+-------------- |  |
| :---: | ---: |
| 0.02 | 24.6629625622 |
| 0.05 | 18.5215370791 |
| 0.07 | 15.7314308304 |
| 0.10 | 14.1887205898 |
| 0.12 | 13.2642785681 |
| 0.15 | 12.6892217095 |
| 0.17 | 12.3276628470 |
| 0.20 | 12.1036117615 |
| 0.22 | 11.9717621751 |
| 0.25 | 11.9035962620 |
| 0.27 | 11.8806036817 |
| 0.30 | 11.8903233353 |
| 0.32 | 11.9241681744 |
| 0.35 | 11.9761582713 |
| 0.37 | 12.0419641658 |
| 0.40 | 12.1185300471 |
| 0.42 | 12.2035857405 |
| 0.45 | 12.2954848038 |
| 0.47 | 12.3930296221 |
| 0.50 | 12.4953627496 |

Lowest SSQ is for $\mathrm{F}=\quad 0.277$

No of years for separable analysis : 8
Age range in the analysis : 1 . . . 9
Year range in the analysis : 1958 . . . 2006
Number of indices of SSB : 0
Number of age-structured indices : 1
Parameters to estimate : 38
Number of observations : 216
Conventional single selection vector model to be fitted.
Survey weighting to be Manual (recommended) or Iterative (M/I) ?-->M
Enter weight for FLT01:West Scotland Summer Acoustic Surv at age 1--> 0.100000000000000
Enter weight for FLT01:West Scotland Summer Acoustic Surv at age 2--> 1.000000000000000
Enter weight for FLT01:West Scotland Summer Acoustic Surv at age 3--> 1.000000000000000
Enter weight for FLT01:West Scotland Summer Acoustic Surv at age 4--> 1.000000000000000
Enter weight for FLT01:West Scotland Summer Acoustic Surv at age 5--> 1.000000000000000
Enter weight for FLT01:West Scotland Summer Acoustic Surv at age 6--> 1.000000000000000
Enter weight for FLT01:West Scotland Summer Acoustic Surv at age 7--> 1.000000000000000
Enter weight for FLT01:West Scotland Summer Acoustic Surv at age 8--> 1.000000000000000
Enter weight for FLT01:West Scotland Summer Acoustic Surv at age 9--> 1.000000000000000
Enter estimates of the extent to which errors
in the age-structured indices are correlated
across ages. This can be in the range 0 (independence)
to 1 (correlated errors).
Enter value for FLT01:West Scotland Summer Acoustic Surv--> 1.000000000000000
Do you want to shrink the final fishing mortality (Y/N) ?-->N
Seeking solution. Please wait.
Aged index weights
FLT01:West Scotland Summer Acoustic Surv


Wts : 0.011 0.111 0.111 0.111 0.1110 .111 0.111 0.111 0.111
F in 2006 at age 4 is 0.238034 in iteration 1
Detailed, Normal or Summary output (D/N/S)-->D
Output page width in characters (e.g. 80..132) ?--> 132
Estimate historical assessment uncertainty ?-->y
Sample from Covariances or Bayes MCMC (C/B) ?-->C
Use default percentiles (Y/N) ?-->y
How many samples to take ?--> 1000
Enter SSB reference level (e.g. MBAL, Bpa..) [t]--> 5.0000000000000000E+04
Succesful exit from ICA

Table 5.6.2. Herring in VIa (N). Catch number at age (millions). N.B. In this table "age" refers to number of rings (winter rings in the otolith).

|  | Herring VIa (north) (run: ICAPGF08/I08) Catch in Number |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 |
| 1 | 15.62 | 53.09 | 3.56 | 13.08 | 55.05 | 11.80 | 26.55 | 299.48 | 211.68 | 207.95 | 220.25 | 37.71 | 238.23 | 207.71 | 534.96 |
| 2 | 30.98 | 67.97 | 102.12 | 45.20 | 92.81 | 78.25 | 82.61 | 19.77 | 500.85 | 27.42 | 94.44 | 92.56 | 99.01 | 335.08 | 621.50 |
| 3 | 145.39 | 35.26 | 60.29 | 61.62 | 22.28 | 53.45 | 70.08 | 62.64 | 33.46 | 218.69 | 21.00 | 71.91 | 253.72 | 412.82 | 175.14 |
| 4 | 39.07 | 116.39 | 22.78 | 33.13 | 67.45 | 11.86 | 26.68 | 59.38 | 60.50 | 37.07 | 159.12 | 23.31 | 111.90 | 302.21 | 54.20 |
| 5 | 24.91 | 24.95 | 48.88 | 22.50 | 44.36 | 40.52 | 7.28 | 22.27 | 40.91 | 39.25 | 13.99 | 211.24 | 27.74 | 101.96 | 66.71 |
| 6 | 27.63 | 17.33 | 11.63 | 12.41 | 19.76 | 26.17 | 24.23 | 5.12 | 19.34 | 29.79 | 23.58 | 21.01 | 142.40 | 25.56 | 25.72 |
| 7 | 17.41 | 17.00 | 10.35 | 5.34 | 24.14 | 8.69 | 18.64 | 22.89 | 5.56 | 11.77 | 15.68 | 42.76 | 21.61 | 154.42 | 10.34 |
| 8 | 9.86 | 7.37 | 6.35 | 4.81 | 6.15 | 13.66 | 8.80 | 18.93 | 17.81 | 5.53 | 6.38 | 26.03 | 27.07 | 16.82 | 55.76 |
| 9 | 7.16 | 8.60 | 4.62 | 2.58 | 7.08 | 6.09 | 15.10 | 19.53 | 27.08 | 25.80 | 10.81 | 26.21 | 24.08 | 32.00 | 16.63 |
| AGE | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 1 | 51.17 | 309.02 | 172.88 | 69.05 | 34.84 | 22.52 | 0.25 | 2.69 | 36.74 | 13.30 | 81.92 | 2.21 | 40.79 | 33.77 | 19.46 |
| 2 | 235.63 | 124.94 | 202.09 | 319.60 | 47.74 | 46.28 | 0.14 | 0.28 | 77.96 | 250.01 | 77.81 | 188.78 | 68.84 | 154.96 | 65.95 |
| 3 | 808.27 | 151.03 | 89.07 | 101.55 | 95.83 | 20.59 | 0.08 | 0.10 | 105.60 | 72.18 | 92.74 | 49.83 | 148.40 | 86.07 | 45.46 |
| 4 | 131.48 | 519.18 | 63.70 | 35.50 | 22.12 | 40.69 | 0.02 | 0.05 | 61.34 | 93.54 | 29.26 | 35.00 | 17.21 | 118.86 | 32.02 |
| 5 | 63.07 | 82.47 | 188.20 | 25.20 | 10.08 | 6.88 | 0.01 | 0.01 | 21.47 | 58.45 | 42.53 | 14.95 | 15.21 | 18.84 | 50.12 |
| 6 | 54.64 | 49.68 | 30.60 | 76.29 | 12.21 | 3.83 | 0.01 | 0.01 | 12.62 | 23.58 | 27.32 | 11.37 | 6.63 | 18.00 | 8.43 |
| 7 | 18.24 | 34.63 | 12.30 | 10.92 | 20.99 | 2.10 | 0.00 | 0.01 | 11.58 | 11.52 | 14.71 | 9.30 | 6.91 | 2.58 | 7.31 |
| 8 | 6.51 | 22.47 | 13.12 | 3.91 | 2.76 | 6.28 | 0.00 | 0.00 | 1.31 | 13.81 | 8.44 | 4.43 | 3.32 | 1.43 | 3.51 |
| 9 | 32.22 | 21.04 | 13.70 | 12.01 | 1.49 | 1.54 | 0.00 | 0.00 | 1.33 | 4.03 | 8.48 | 1.96 | 2.19 | 1.97 | 5.98 |
| AGE | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 1 | 1.71 | 6.22 | 14.29 | 26.40 | 5.25 | 17.72 | 1.73 | 0.27 | 1.95 | 1.19 | 9.09 | 7.63 | 3.57 | 0.14 | 0.99 |
| 2 | 119.38 | 36.76 | 40.87 | 23.01 | 24.47 | 95.29 | 36.55 | 82.18 | 37.85 | 55.81 | 74.17 | 35.25 | 18.16 | 81.03 | 38.48 |
| 3 | 41.73 | 109.50 | 40.78 | 25.23 | 24.92 | 18.71 | 40.19 | 30.40 | 30.90 | 34.97 | 34.57 | 93.91 | 17.26 | 14.94 | 93.98 |
| 4 | 28.42 | 18.92 | 74.28 | 28.21 | 23.73 | 10.98 | 6.01 | 21.27 | 9.22 | 31.66 | 31.91 | 25.08 | 40.67 | 9.31 | 9.01 |
| 5 | 19.76 | 18.11 | 26.52 | 37.52 | 21.82 | 13.27 | 7.43 | 5.38 | 7.51 | 23.12 | 22.87 | 13.36 | 12.26 | 24.48 | 18.11 |
| 6 | 28.55 | 7.59 | 13.30 | 13.53 | 33.87 | 14.80 | 8.10 | 4.21 | 2.50 | 17.50 | 14.37 | 7.53 | 7.12 | 9.28 | 28.02 |
| 7 | 3.25 | 15.01 | 9.88 | 7.58 | 6.35 | 19.19 | 10.52 | 8.80 | 4.70 | 10.33 | 8.64 | 3.25 | 3.08 | 6.62 | 9.04 |
| 8 | 2.22 | 1.62 | 21.46 | 6.89 | 4.32 | 4.71 | 12.16 | 7.97 | 8.46 | 5.21 | 2.83 | 1.26 | 1.45 | 4.61 | 1.55 |
| 9 | 2.36 | 3.50 | 5.52 | 4.46 | 5.51 | 3.74 | 10.21 | 9.79 | 31.11 | 9.88 | 3.33 | 1.09 | 0.46 | 1.00 | 1.42 |

Table 5.6.2. Herring in VIa (N). Catch number at age (millions). Continued
Catch in Number

| AGE | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 0.06 | 0.00 | 0.18 | 0.13 |
| 2 | 33.33 | 6.84 | 9.63 | 6.69 |
| 3 | 46.87 | 22.22 | 23.24 | 9.19 |
| 4 | 53.77 | 27.82 | 20.60 | 13.64 |
| 5 | 7.46 | 45.78 | 10.24 | 41.07 |
| 6 | 4.34 | 3.92 | 9.78 | 27.78 |
| 7 | 12.82 | 7.64 | 1.01 | 20.97 |
| 8 | 9.19 | 8.48 | 1.19 | 3.04 |
| 9 | 1.41 | 4.01 | 1.43 | 5.09 |

Table 5.6.3. Herring in VIa (N). Weight in the catch (kg). N.B. In this table "age" refers to number of rings (winter rings in the otolith).
Weights at age in the catches ( Kg )

| AGE | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 | 0.07900 |
| 2 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.10400 | 0.1040 | 0.10400 |
| 3 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 | 0.13000 |
| 4 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 | 0.15800 |
| 5 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 |
| 6 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 | 0.17000 |
| 7 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 | 0.18000 |
|  | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 | 0.18300 |
| 9 | 0.18500 | 0.18500 | 0.18500 | 0.18500 | 0.18500 | 0.18500 | 0.1850 | 0.1850 | 0.1850 | 0.18500 | 0.1850 | 0.1850 | 0.1850 | 0.185 | - 18 |


| AGE | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


$2 \quad 0.090000 .090000 .090000 .090000 .090000 .090000 .09000$ 0.09000 0.09000 0.08000 0.08000 0.08000 0.06900 0.11300 0.0730 ( .121000 .121000 .121000 .121000 .121000 .121000 .121000 .121000 .121000 .140000 .140000 .140000 .103000 .145000 .14300 $3 \quad 0.158000 .158000 .158000 .158000 .158000 .158000 .158000 .158000 .158000 .175000 .175000 .175000 .134000 .173000 .18300$ 0.175000 .175000 .175000 .175000 .175000 .175000 .175000 .175000 .175000 .205000 .205000 .205000 .161000 .196000 .21100 0.186000 .186000 .186000 .186000 .186000 .186000 .186000 .186000 .186000 .231000 .231000 .231000 .182000 .215000 .22000 0.20600 0. 206000.206000 .206000 .206000 .206000 .206000 .206000 .206000 .253000 .253000 .253000 .199000 .230000 .23800 0.218000 .218000 .218000 .218000 .218000 .218000 .218000 .218000 .218000 .270000 .270000 .270000 .213000 .242000 .24100



## Table 5.6.3. Herring in VIa (N). Weight in the catch (kg). Continued

Weights at age in the catches (Kg)

| AGE | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.08000 | 0.08200 | 0.07900 | 0.08400 | 0.09100 | 0.08900 | 0.08300 | 0.10600 | 0.08100 | 0.08900 | 0.09700 | 0.07600 | 0.08340 | 0.04900 | 0.10700 |
| 2 | 0.11200 | 0.14200 | 0.12900 | 0.11800 | 0.11900 | 0.12800 | 0.14200 | 0.14200 | 0.13400 | 0.13600 | 0.13800 | 0.13000 | 0.13730 | 0.14000 | 0.14600 |
| 3 | 0.15700 | 0.14500 | 0.17300 | 0.16000 | 0.18300 | 0.15800 | 0.16700 | 0.18100 | 0.17800 | 0.17700 | 0.15900 | 0.15800 | 0.16370 | 0.16300 | 0.16300 |
| 4 | 0.17700 | 0.19100 | 0.18200 | 0.20300 | 0.19600 | 0.19700 | 0.19000 | 0.19100 | 0.21000 | 0.20500 | 0.18200 | 0.17500 | 0.18290 | 0.18300 | 0.17300 |
| 5 | 0.20300 | 0.19000 | 0.20900 | 0.21100 | 0.22700 | 0.20600 | 0.19500 | 0.19800 | 0.23000 | 0.22200 | 0.19900 | 0.19100 | 0.20140 | 0.19200 | 0.16000 |
| 6 | 0.19400 | 0.21300 | 0.22400 | 0.22900 | 0.21900 | 0.22800 | 0.20100 | 0.21400 | 0.23300 | 0.22300 | 0.21800 | 0.21000 | 0.21470 | 0.19600 | 0.17900 |
| 7 | 0.24000 | 0.21600 | 0.22800 | 0.23600 | 0.24400 | 0.22300 | 0.24400 | 0.20800 | 0.26200 | 0.21900 | 0.22700 | 0.22500 | 0.23940 | 0.20500 | 0.18700 |
| 8 | 0.21300 | 0.20400 | 0.23700 | 0.26100 | 0.25600 | 0.26200 | 0.23400 | 0.22700 | 0.24700 | 0.23800 | 0.21200 | 0.22300 | 0.28120 | 0.22500 | 0.24500 |
| 9 | 0.22800 | 0.24300 | 0.24700 | 0.27100 | 0.25600 | 0.26300 | 0.26600 | 0.27700 | 0.29100 | 0.26300 | 0.19900 | 0.22600 | 0.25260 | 0.27200 | 0.28100 |


| AGE | 2003 | 2004 | 2005 | 20 |
| :---: | :---: | :---: | :---: | :---: |

0.060000 .000000 .108400 .09080 0.145000 .154000 .132700 .15800 0.160000 .173000 .163200 .16760 0.169000 .195000 .184500 .19290 0.186000 .216000 .210800 .20760 0.200000 .220000 .225800 .22510 $\begin{array}{llll}0.20000 & 0.22000 & 0.22580 & 0.22510 \\ 0.19400 & 0.19900 & 0.23410 & 0.24430\end{array}$ $\begin{array}{llll}0.18600 & 0.19000 & 0.25560 & 0.26150\end{array}$
$0.29400 \quad 0.31100 \quad 0.24960 \quad 0.27500$
Table 5.6.4. Herring in VIa (N). Weight in the stock (kg). N.B. In this table "age" refers to number of rings (winter rings in the otolith).

```
Weights at age in the stock ( Kg )
```

| AGE | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.09000 0.164000 .164000 .164000 .164000 .164000 .164000 .164000 .164000 .164000 .164000 .164000 .164000 .164000 .164000 .16400 0.208000 .208000 .208000 .208000 .208000 .208000 .208000 .208000 .208000 .208000 .208000 .208000 .208000 .208000 .20800 0.233000 .233000 .233000 .233000 .233000 .233000 .233000 .233000 .233000 .233000 .233000 .233000 .233000 .233000 .23300
 0.252000 .252000 .252000 .252000 .252000 .252000 .252000 .252000 .252000 .252000 .252000 .252000 .252000 .252000 .25200 0.25800 0. 25800 0. 25800 0. 25800 0. 25800 0. 25800 0. 25800 0. 25800 0. 25800 0. 25800 0. 25800 0. 25800 0. 25800 0. $25800 \quad 0.25800$ 0.26900 0. 26900 0. 26900 0. 26900 0. 269000.269000 .269000 .269000 .269000 .269000 .269000 .269000 .269000 .269000 .26900 0.29200 0. 29200 0. 29200 0. 29200 0. 29200 0. 29200 0. 29200 0. 29200 0. 29200 0. 29200 0. 29200 0. 29200 0. 29200 0. $29200 \quad 0.29200$

Table 5.6.4. Herring in VIa (N). Continued.

Weights at age in the stock $(\mathrm{Kg})$

| AGE | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.09000 | 0.0900 |
| 2 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0.16400 | 0. |
| 3 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 | 0.20800 |
| 4 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0.23300 | 0. |
| 5 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0.24600 | 0. |
| 6 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0.25200 | 0. |
| 7 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | - |
| 8 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0.26900 | 0. |
| 9 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0.29200 | 0. |


| AGE | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

 $\begin{array}{lllllllllllll}0.16400 & 0.16400 & 0.16400 & 0.16400 & 0.16400 & 0.16200 & 0.15000 & 0.14400 & 0.14000 & 0.15000 & 0.13800 & 0.13700 & 0.14100 \\ 0.13200 & 0.15300\end{array}$ 0.208000 .208000 .208000 .208000 .208000 .196000 .192000 .191000 .180000 .189000 .176000 .166000 .173000 .170000 .17700



 0.292000 .292000 .292000 .292000 .292000 .276000 .296000 .293000 .263000 .287000 .249000 .245000 .230000 .254000 .25900

| AGE | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 0.06400 | 0.05900 | 0.07510 | 0.07500 |
| 2 | 0.13800 | 0.13800 | 0.12960 | 0.13500 |
| 3 | 0.17600 | 0.15900 | 0.15380 | 0.16600 |
| 4 | 0.19000 | 0.18000 | 0.16650 | 0.18500 |
| 5 | 0.20400 | 0.18900 | 0.18020 | 0.19200 |
| 6 | 0.21300 | 0.20200 | 0.19110 | 0.20400 |
| 7 | 0.21700 | 0.21300 | 0.21250 | 0.21100 |
| 8 | 0.22300 | 0.21400 | 0.20300 | 0.22400 |
| 9 | 0.22800 | 0.20600 | 0.22840 | 0.23100 |

Table 5.6.5. Herring in VIa (N). Natural mortality. N.B. In this table "age" refers to number of rings (winter rings in the otolith).
Natural Mortality (per year)

| AGE | 1957 | 1958 | 1959 | 1960 | 1961 |  | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |  | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |  | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |  | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |  | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |  | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |  | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |  | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |  | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 9 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |  | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |

Table 5.6.6. Herring in VIa ( N ). Proportion mature. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| AGE | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 |
| 3 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| AGE | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.5700 |
| 3 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 0.9600 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Table 5.6.6. Herring in VIa (N). Proportion mature. Continued

Proportion of fish spawning

| AGE | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 200 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 | 0.5700 | 0.5700 | 0.5700 | 0.5700 | 0.4700 | 0.9300 | 0.4800 | 0.1900 | 0.7600 | 0.5700 | 0.8500 | 0.5700 | 0.4500 | 0.9300 | 0.9200 |
| 3 | 0.9600 | 0.9600 | 0.9600 | 0.9600 | 1.0000 | 0.9600 | 0.9200 | 0.9800 | 0.9400 | 0.9600 | 0.9700 | 0.9800 | 0.9200 | 0.9900 | 1.0000 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.000 |

$\left.\begin{array}{c|cccr}\hline \text { AGE } & \text { | } & 2003 & 2004 & 2005\end{array}\right) 2006$

Table 5.6.7. Herring in VIa (N). Tuning indices. N.B. In this table "age" refers to number of rings (winter rings in the otolith)
AGE-STRUCTURED INDICES

| AGE | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 249.1 |  |  |  | 338.3 | 74.3 | 2.8 | 494.2 | 460.6 | 41.2 | 792.3 | 1221.7 | 534.2 | 447.6 | 313.1 |
| 2 | 578.4 |  |  | **** | 294.5 | 503.4 | 750.3 | 542.1 | 1085.1 | 576.5 | 641.9 | 794.6 | 322.4 | 316.2 | 1062.0 |
| 3 | 551.1 |  |  | ******* | 327.9 | 211.0 | 681.2 | 607.7 | 472.7 | 802.5 | 286.2 | 666.8 | 1388.0 | 337.1 | 217.7 |
| 4 | 353.1 |  |  | ******* | 367.8 | 258.1 | 653.0 | 285.6 | 450.2 | 329.1 | 167.0 | 471.1 | 432.0 | 899.5 | 172.8 |
| 5 | 752.6 |  |  |  | 488.3 | 414.8 | 544.0 | 306.8 | 153.0 | 95.4 | 66.1 | 179.1 | 308.0 | 393.4 | 437.5 |
| 6 | 111.6 |  |  |  | 176.3 | 240.1 | 865.2 | 268.1 | 187.1 | 60.6 | 49.5 | 79.3 | 138.7 | 247.6 | 132.6 |
| 7 | 48.1 |  |  |  | 98.7 | 105.7 | 284.1 | 406.8 | 169.2 | 77.4 | 16.3 | 28.1 | 86.5 | 199.5 | 102.8 |
| 8 | 15.9 |  |  |  | 89.8 | 56.7 | 151.7 | 173.7 | 236.6 | 78.2 | 29.0 | 13.8 | 27.6 | 95.0 | 52.4 |
| 9 | 6.5 |  |  |  | 58.0 | 63.4 | 156.2 | 131.9 | 201.5 | 114.8 | 24.4 | 36.8 | 35.4 | 65.0 | 34.7 |

$\times 10 \wedge 3$

Table 5.6.7. Herring in VIa ( N ). Tuning indices. Continued

| AGE | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 424.7 | 438.8 | 564.0 | 50.2 | 112.3 |
| 2 | 436.0 | 1039.4 | 274.5 | 243.4 | 835.2 |
| 3 | 1436.9 | 932.5 | 760.2 | 230.3 | 387.9 |
| 4 | 199.8 | 1471.8 | 442.3 | 423.1 | 284.5 |
| 5 | 161.7 | 181.3 | 577.2 | 245.1 | 582.2 |
| 6 | 424.3 | 129.2 | 55.7 | 152.8 | 414.7 |
| 7 | 152.3 | 346.7 | 61.8 | 12.6 | 227.0 |
| 8 | 67.5 | 114.3 | 82.2 | 39.0 | 21.7 |
| 9 | 59.5 | 75.2 | 76.3 | 26.8 | 59.3 |

Table 5.6.8. Herring in VIa ( $\mathbf{N}$ ). Weighting factors for the catch in numbers. N.B. In this table "age" refers to number of rings (winter rings in the otolith). Weighting factors for the catches in number

| AGE | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 2 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 3 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Table 5.6.9. Herring in VIa (N). Fishing mortality (per year). N.B. In this table "age" refers to number of rings (winter rings in the otolith).


Table 5.6.9. Herring in VIa (N). Fishing mortality (per year). Continued.
Fishing Mortality (per year)

| AGE | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0017 | 0.0015 | 0.0010 | 0.0021 |
| 2 | 0.1437 | 0.1257 | 0.0794 | 0.1731 |
| 3 | 0.2447 | 0.2141 | 0.1352 | 0.2947 |
| 4 | 0.1977 | 0.1729 | 0.1092 | 0.2380 |
| 5 | 0.2632 | 0.2303 | 0.1454 | 0.3170 |
| 6 | 0.2095 | 0.1833 | 0.1157 | 0.2523 |
| 7 | 0.2067 | 0.1809 | 0.1142 | 0.2490 |
| 8 | 0.1977 | 0.1729 | 0.1092 | 0.2380 |
| 9 | 0.1977 | 0.1729 | 0.1092 | 0.2380 |

Table 5.6.10. Herring in VIa (N). Population abundance (1 January, millions). N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Population Abundance (1 January) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 |
| 1 | 2212.7 | 2207.1 | 647.5 | 1317.9 | 2357.3 | 2154.7 | 993.2 | 7931.8 | 1073.2 | 2509.7 | 4107.2 | 3000.7 | 3441.3 | 9575.7 | 2676.2 |
| 2 | 415.9 | 804.9 | 781.2 | 236.1 | 477.2 | 835.3 | 785.8 | 350.0 | 2744.5 | 275.2 | 803.6 | 1383.7 | 1082.0 | 1128.6 | 3402.2 |
| 3 | 636.7 | 281.6 | 538.2 | 491.5 | 136.4 | 274.5 | 551.9 | 511.5 | 242.4 | 1606.2 | 180.4 | 514.7 | 945.8 | 716.9 | 551.7 |
| 4 | 144.8 | 390.6 | 198.8 | 386.3 | 346.9 | 91.6 | 176.6 | 388.7 | 362.4 | 168.3 | 1118.1 | 128.8 | 356.6 | 546.5 | 220.0 |
| 5 | 109.4 | 94.0 | 243.1 | 158.2 | 318.0 | 249.8 | 71.6 | 134.5 | 295.3 | 270.4 | 117.1 | 860.6 | 94.4 | 216.6 | 209.2 |
| 6 | 98.4 | 75.4 | 61.4 | 173.6 | 121.8 | 245.7 | 187.6 | 57.9 | 100.5 | 228.4 | 207.4 | 92.7 | 578.3 | 59.1 | 99.6 |
| 7 | 57.8 | 62.8 | 51.8 | 44.5 | 145.3 | 91.5 | 197.4 | 146.7 | 47.5 | 72.6 | 178.4 | 165.3 | 63.9 | 388.2 | 29.3 |
| 8 | 44.5 | 35.8 | 40.7 | 37.0 | 35.2 | 108.5 | 74.5 | 160.9 | 111.0 | 37.7 | 54.5 | 146.5 | 109.0 | 37.4 | 205.1 |
| 9 | 32.3 | 41.8 | 29.6 | 19.9 | 40.6 | 48.4 | 127.9 | 166.1 | 168.8 | 175.9 | 92.5 | 147.5 | 97.0 | 71.1 | 61.2 |
| AGE | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 1 | 1075.1 | 1675.0 | 2116.7 | 614.5 | 626.1 | 916.0 | 1219.4 | 892.1 | 1667.5 | 777.7 | 3036.2 | 1152.7 | 1214.2 | 904.6 | 2148.7 |
| 2 | 682.3 | 365.9 | 441.2 | 679.2 | 186.5 | 210.2 | 323.9 | 448.5 | 326.6 | 592.1 | 278.4 | 1069.4 | 422.8 | 423.1 | 313.2 |
| 3 | 1990.6 | 305.9 | 165.2 | 156.6 | 234.2 | 97.6 | 116.3 | 239.8 | 332.0 | 175.7 | 227.7 | 140.1 | 631.3 | 254.5 | 182.3 |
| 4 | 294.6 | 906.7 | 115.7 | 56.0 | 38.3 | 106.0 | 61.4 | 95.2 | 196.3 | 177.1 | 79.3 | 103.5 | 70.1 | 383.5 | 131.2 |
| 5 | 147.6 | 142.2 | 330.4 | 44.6 | 17.2 | 13.8 | 57.4 | 55.5 | 86.1 | 119.5 | 71.9 | 44.0 | 60.5 | 47.1 | 234.3 |
| 6 | 126.1 | 73.9 | 50.9 | 121.3 | 16.6 | 6.0 | 6.0 | 51.9 | 50.2 | 57.5 | 52.8 | 24.9 | 25.7 | 40.3 | 24.8 |
| 7 | 65.7 | 62.4 | 20.1 | 17.2 | 37.9 | 3.5 | 1.9 | 5.4 | 47.0 | 33.5 | 29.7 | 22.0 | 11.8 | 16.9 | 19.4 |
| 8 | 16.7 | 42.2 | 23.7 | 6.6 | 5.2 | 14.4 | 1.2 | 1.7 | 4.9 | 31.5 | 19.4 | 13.0 | 11.1 | 4.1 | 12.9 |
| 9 | 82.8 | 39.5 | 24.8 | 20.2 | 2.8 | 3.6 | 8.9 | 9.1 | 4.9 | 9.2 | 19.5 | 5.8 | 7.3 | 5.7 | 21.9 |

$\times 10 \wedge 6$

Table 5.6.10. Herring in VIa (N). Population abundance (1 January, millions). Continued.

| Population Abundance (1 January) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 1 | 927.0 | 880.1 | 438.5 | 384.7 | 799.7 | 602.4 | 855.3 | 668.1 | 866.0 | 1538.7 | 482.4 | 307.4 | 1726.8 | 955.6 | 921.2 |
| 2 | 779.2 | 340.0 | 320.2 | 153.0 | 126.3 | 291.1 | 211.3 | 313.7 | 245.6 | 317.4 | 565.4 | 172.2 | 112.8 | 634.3 | 350.9 |
| 3 | 175.9 | 475.3 | 220.5 | 202.3 | 93.7 | 72.7 | 134.9 | 125.4 | 162.5 | 149.7 | 187.6 | 355.5 | 105.1 | 73.4 | 406.6 |
| 4 | 108.4 | 106.5 | 290.7 | 143.8 | 142.9 | 54.3 | 42.7 | 74.4 | 75.3 | 105.3 | 91.1 | 122.5 | 209.2 | 69.0 | 47.0 |
| 5 | 88.3 | 71.1 | 78.4 | 192.6 | 103.4 | 106.8 | 38.8 | 33.0 | 47.1 | 59.4 | 65.2 | 52.2 | 84.9 | 158.3 | 51.1 |
| 6 | 164.5 | 61.2 | 47.2 | 45.8 | 138.7 | 72.8 | 84.0 | 28.0 | 24.7 | 35.5 | 31.9 | 37.4 | 33.1 | 60.5 | 109.9 |
| 7 | 14.4 | 121.7 | 48.1 | 30.1 | 28.6 | 93.4 | 51.9 | 68.3 | 21.4 | 20.0 | 15.6 | 15.2 | 25.5 | 24.8 | 44.4 |
| 8 | 10.7 | 10.7 | 95.9 | 34.2 | 20.0 | 19.9 | 66.3 | 36.9 | 53.4 | 14.9 | 8.3 | 6.0 | 10.4 | 19.1 | 18.2 |
| 9 | 11.3 | 21.6 | 24.7 | 22.1 | 25.6 | 15.8 | 55.6 | 45.4 | 196.5 | 28.2 | 9.8 | 4.9 | 2.9 | 5.8 | 7.7 |


| AGE | 2003 | 2004 | 2005 | 2006 | 2007 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 363.6 | 259.4 | 168.2 | 112.2 | 400.6 |
| 2 | 338.3 | 133.5 | 95.3 | 61.8 | 41.2 |
| 3 | 222.5 | 217.0 | 87.2 | 65.2 | 38.5 |
| 4 | 255.4 | 142.6 | 143.4 | 62.4 | 39.8 |
| 5 | 34.3 | 189.7 | 108.6 | 116.4 | 44.5 |
| 6 | 34.8 | 23.9 | 136.3 | 84.9 | 76.7 |
| 7 | 79.3 | 25.5 | 18.0 | 109.9 | 59.7 |
| 8 | 32.1 | 58.3 | 19.3 | 14.5 | 77.5 |
| 9 | 8.2 | 26.5 | 14.5 | 25.2 | 28.3 |

$\times 10 \wedge 6$
Table 5.6.11. Herring in VIa (N). Predicted catch in number. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Predicted Catch in Number |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 453. | 1703. | 1050. | 1089. | 397. | 248. | 102. | 148. |
| 2 | 26365. | 11899. | 74071. | 43848. | 39264. | 13674. | 6295. | 8526. |
| 3 | 91161. | 18944. | 14586. | 86129. | 43965. | 38051. | 10019. | 15159. |
| 4 | 27357. | 32575. | 11858. | 8625. | 43665. | 21587. | 14131. | 12601. |
| 5 | 14905. | 17114. | 35145. | 12093. | 7576. | 37200. | 13998. | 30165. |
| 6 | 8781. | 5439. | 10974. | 21262. | 6272. | 3810. | 14191. | 18066. |
| 7 | 3542. | 4134. | 4440. | 8480. | 14114. | 4027. | 1848. | 23090. |
| 8 | 1331. | 1625. | 3289. | 3345. | 5487. | 8828. | 1900. | 2931. |

$x 10 \wedge 6$

Table 5.6.12. Herring in VIa (N). Predicted index values. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| AGE | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 805.6 |  | ***** | ******* | 136.6 | 300.5 | 221.8 | 322.6 | 252.4 | 326.6 | 581.0 | 179.3 | 116.0 | 651.9 | 360.8 |
| 2 | 747.8 | ***** | ***** | * | 383.0 | 305.7 | 626.0 | 519.9 | 716.8 | 613.1 | 779.3 | 1435.7 | 430.1 | 291.9 | 1627.7 |
| 3 | 620.6 |  | * | *** | 756.3 | 314.7 | 245.9 | 440.7 | 429.2 | 579.6 | 516.0 | 671.9 | 1203.1 | 377.4 | 260.1 |
| 4 | 487.3 |  | * | *** | 553.6 | 561.7 | 208.3 | 170.9 | 267.3 | 305.0 | 373.7 | 309.9 | 462.1 | 828.1 | 270.0 |
| 5 | 822.5 | * | * | ** | 685.5 | 363.5 | 398.7 | 138.2 | 119.9 | 172.0 | 180.1 | 204.9 | 173.4 | 300.5 | 552.4 |
| 6 | 75.4 | ***** | ***** | * | 144.8 | 456.6 | 247.2 | 306.5 | 98.7 | 89.9 | 92.7 | 87.1 | 123.8 | 115.5 | 208.7 |
| 7 | 51.4 |  |  | * | 88.5 | 86.2 | 284.5 | 158.3 | 219.5 | 64.4 | 45.5 | 33.9 | 45.5 | 80.0 | 77.0 |
| 8 | 34.2 |  | * | ** | 95.8 | 55.6 | 54.3 | 188.2 | 102.5 | 154.4 | 37.0 | 20.9 | 16.4 | 30.2 | 54.7 |
| 9 | 66.2 |  |  |  | 70.4 | 80.7 | 49.0 | 179.6 | 143.1 | 645.8 | 79.7 | 28.0 | 15.3 | 9.6 | 18.9 |


| AGE | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 347.7 | 137.3 | 97.9 | 63.5 | 42.4 |
| 2 | 895.3 | 868.5 | 346.2 | 253.4 | 156.2 |
| 3 | 1426.1 | 789.1 | 782.6 | 328.4 | 225.0 |
| 4 | 182.4 | 1000.8 | 566.5 | 589.8 | 239.1 |
| 5 | 176.5 | 119.9 | 674.3 | 404.3 | 394.6 |
| 6 | 375.7 | 120.1 | 83.5 | 495.0 | 286.3 |
| 7 | 136.6 | 246.4 | 80.5 | 58.8 | 333.6 |
| 8 | 51.7 | 91.9 | 169.4 | 58.0 | 40.7 |
| 9 | 25.0 | 26.8 | 87.4 | 49.6 | 80.3 |

$x 10 \wedge 3$
Table 5.6.13. Herring in VIa (N). Fitted selection pattern. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Fitted Selection Pattern |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 |
| 1 | 0.0338 | 0.1034 | 0.0681 | 0.1675 | 0.1645 | 0.0595 | 0.2494 | 0.3507 | 1.8746 | 0.5285 | 0.5439 | 0.0952 | 0.2882 | 0.0405 | 1.2271 |
| 2 | 0.2707 | 0.2741 | 1.2745 | 2.6357 | 1.1099 | 0.7837 | 0.7492 | 0.3863 | 1.2240 | 0.4651 | 0.9003 | 0.3819 | 0.2801 | 0.4832 | 0.7897 |
| 3 | 0.8687 | 0.3963 | 1.0270 | 1.5736 | 0.8679 | 1.6495 | 0.8722 | 0.8288 | 0.8555 | 0.6180 | 0.8476 | 0.7926 | 0.8746 | 1.1408 | 1.4301 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.8205 | 0.8710 | 1.8478 | 1.7127 | 0.6937 | 1.2774 | 0.6540 | 1.0923 | 0.8156 | 0.6291 | 0.8282 | 1.4125 | 0.9235 | 0.7869 | 1.3602 |
| 6 | 1.0487 | 0.7370 | 1.7285 | 0.8272 | 0.8178 | 0.8123 | 0.8439 | 0.5578 | 1.1704 | 0.5607 | 0.7856 | 1.2888 | 0.7492 | 0.6992 | 1.0560 |
| 7 | 1.1399 | 0.8909 | 1.8349 | 1.4277 | 0.8398 | 0.7195 | 0.6048 | 1.0232 | 0.6807 | 0.7098 | 0.5986 | 1.5016 | 1.0964 | 0.6254 | 1.5427 |
| 8 | 0.7948 | 0.6497 | 1.3924 | 1.5545 | 0.8871 | 0.9705 | 0.7667 | 0.7542 | 0.9569 | 0.6364 | 0.8097 | 0.9795 | 0.7567 | 0.7390 | 1.1222 |
| 9 | 0.7948 | 0.6497 | 1.3924 | 1.5545 | 0.8871 | 0.9705 | 0.7667 | 0.7542 | 0.9569 | 0.6364 | 0.8097 | 0.9795 | 0.7567 | 0.7390 | 1.1222 |

Table 5.6.13. Herring in VIa (N). Fitted selection pattern. Continued.
Fitted Selection Pattern

| AGE | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.1238 | 0.3674 | 0.1601 | 0.1781 | 0.0991 | 0.0770 | 0.9852 | 8.4923 | 0.0892 | 0.0341 | 0.0890 | 0.0069 | 0.1826 | 0.1543 | 0.0488 |
| 2 | 0.7992 | 0.5443 | 0.8612 | 0.7079 | 0.3771 | 0.5680 | 1.5602 | 1.2790 | 0.8078 | 0.8178 | 0.7909 | 0.5195 | 0.6976 | 1.3804 | 0.9375 |
| 3 | 0.9333 | 0.8487 | 1.0331 | 1.1188 | 0.6426 | 0.5134 | 2.2456 | 0.7761 | 1.0805 | 0.7433 | 1.2054 | 1.1273 | 1.0028 | 1.1785 | 1.0818 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.9422 | 1.0205 | 1.0560 | 0.8242 | 1.0248 | 1.4352 | 0.7316 | 0.4369 | 0.7645 | 0.8927 | 1.9649 | 1.0054 | 1.0282 | 1.3810 | 0.8591 |
| 6 | 0.9608 | 1.3227 | 1.1551 | 0.9852 | 1.5744 | 2.1014 | 4.3343 | 0.3234 | 0.7711 | 0.6986 | 1.5887 | 1.4824 | 1.0615 | 1.6034 | 1.4895 |
| 7 | 0.5469 | 0.9521 | 1.1907 | 1.0056 | 0.9365 | 1.9048 | 6.9537 | 2.7709 | 0.7541 | 0.5570 | 1.4893 | 1.3342 | 3.1718 | 0.4438 | 1.6923 |
| 8 | 0.8316 | 0.8930 | 1.0066 | 0.8978 | 0.8646 | 1.1795 | 2.7076 | 1.1113 | 0.8343 | 0.7644 | 1.2432 | 1.0089 | 1.2625 | 1.1373 | 1.1385 |
| 9 | 0.8316 | 0.8930 | 1.0066 | 0.8978 | 0.8646 | 1.1795 | 2.7076 | 1.1113 | 0.8343 | 0.7644 | 1.2432 | 1.0089 | 1.2625 | 1.1373 | 1.1385 |
| AGE | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 1 | 0.0091 | 0.0544 | 0.1691 | 0.4940 | 0.0545 | 0.1994 | 0.0200 | 0.0018 | 0.0260 | 0.0032 | 0.0661 | 0.0087 | 0.0087 | 0.0087 | 0.0087 |
| 2 | 0.6044 | 0.6457 | 0.5108 | 0.8260 | 1.3166 | 1.9708 | 1.3908 | 1.0038 | 1.4216 | 0.5974 | 0.3590 | 0.7272 | 0.7272 | 0.7272 | 0.7272 |
| 3 | 0.9393 | 1.4135 | 0.7292 | 0.6411 | 1.8012 | 1.3936 | 2.4765 | 0.8688 | 1.7051 | 0.7833 | 0.4957 | 1.2382 | 1.2382 | 1.2382 | 1.2382 |
|  | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.8322 | 1.5044 | 1.4030 | 0.9920 | 1.3065 | 0.5872 | 1.4068 | 0.5271 | 1.3302 | 1.3814 | 1.0015 | 1.3318 | 1.3318 | 1.3318 | 1.3318 |
| 6 | 0.6258 | 0.6763 | 1.1234 | 1.6077 | 1.5441 | 1.0069 | 0.6688 | 0.4812 | 0.8171 | 1.9098 | 1.3956 | 1.0601 | 1.0601 | 1.0601 | 1.0601 |
| 7 | 0.8387 | 0.6720 | 0.7775 | 1.3314 | 1.3830 | 1.0195 | 1.4978 | 0.4082 | 1.9084 | 2.0503 | 1.8882 | 1.0459 | 1.0459 | 1.0459 | 1.0459 |
| 8 | 0.7678 | 0.9066 | 0.8580 | 1.0313 | 1.3375 | 1.2002 | 1.3386 | 0.7203 | 1.3210 | 1.2089 | 0.9613 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 0.7678 | 0.9066 | 0.8580 | 1.0313 | 1.3375 | 1.2002 | 1.3386 | 0.7203 | 1.3210 | 1.2089 | 0.9613 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |


| AGE | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0087 | 0.0087 | 0.0087 | 0.0087 |
| 2 | 0.7272 | 0.7272 | 0.7272 | 0.7272 |
| 3 | 1.2382 | 1.2382 | 1.2382 | 1.2382 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.3318 | 1.3318 | 1.3318 | 1.3318 |
| 6 | 1.0601 | 1.0601 | 1.0601 | 1.0601 |
| 7 | 1.0459 | 1.0459 | 1.0459 | 1.0459 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 9 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Table 5.6.14. Herring in VIa (N). Stock summary. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

STOCK SUMMARY

| Year | Recruits | Total | Spawning | Landings | Yield | Mean F | SoP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 1 | Biomass | Biomass |  | /SSB | Ages |  |
|  | thousands | tonnes | tonnes | tonnes | ratio | 3-6 | (\%) |
| 1958 | 2212700 | 521599 | 213788 | 59669 | 0.2791 | 0.3105 | 133 |
| 1959 | 2207110 | 560404 | 229855 | 65221 | 0.2837 | 0.2810 | 137 |
| 1960 | 647460 | 452889 | 266482 | 63759 | 0.2393 | 0.1795 | 176 |
| 1961 | 1317900 | 459481 | 266665 | 46353 | 0.1738 | 0.1206 | 171 |
| 1962 | 2357340 | 567345 | 255902 | 58195 | 0.2274 | 0.1927 | 129 |
| 1963 | 2154720 | 599628 | 279650 | 49030 | 0.1753 | 0.1730 | 143 |
| 1964 | 993210 | 547433 | 325431 | 64234 | 0.1974 | 0.1454 | 173 |
| 1965 | 7931830 | 1145548 | 331998 | 68669 | 0.2068 | 0.1519 | 116 |
| 1966 | 1073170 | 870956 | 443967 | 100619 | 0.2266 | 0.1850 | 98 |
| 1967 | 2509730 | 848638 | 473104 | 90400 | 0.1911 | 0.1843 | 123 |
| 1968 | 4107180 | 968245 | 448629 | 84614 | 0.1886 | 0.1400 | 125 |
| 1969 | 3000720 | 994218 | 484835 | 107170 | 0.2210 | 0.2366 | 132 |
| 1970 | 3441280 | 1010076 | 451154 | 165930 | 0.3678 | 0.3534 | 136 |
| 1971 | 9575740 | 1522522 | 322203 | 207167 | 0.6430 | 0.7801 | 98 |
| 1972 | 2676160 | 1121975 | 449079 | 164756 | 0.3669 | 0.3621 | 97 |
| 1973 | 1075080 | 805073 | 388257 | 210270 | 0.5416 | 0.6026 | 95 |
| 1974 | 1674990 | 578215 | 205535 | 178160 | 0.8668 | 0.9531 | 88 |
| 1975 | 2116660 | 437091 | 108162 | 114001 | 1.0540 | 0.9061 | 98 |
| 1976 | 614530 | 265951 | 74599 | 93642 | 1.2553 | 1.0609 | 100 |
| 1977 | 626140 | 164986 | 53217 | 41341 | 0.7768 | 0.9779 | 109 |
| 1978 | 915990 | 172660 | 50015 | 22156 | 0.4430 | 0.6483 | 99 |
| 1979 | 1219400 | 220389 | 76219 | 60 | 0.0008 | 0.0007 | 99 |
| 1980 | 892140 | 257150 | 126093 | 306 | 0.0024 | 0.0004 | 99 |
| 1981 | 1667530 | 367133 | 133595 | 51420 | 0.3849 | 0.3585 | 103 |
| 1982 | 777670 | 308579 | 111549 | 92360 | 0.8280 | 0.6683 | 96 |
| 1983 | 3036180 | 434311 | 82975 | 63523 | 0.7656 | 0.7033 | 97 |
| 1984 | 1152680 | 360340 | 123173 | 56012 | 0.4547 | 0.5044 | 105 |
| 1985 | 1214240 | 355762 | 152361 | 39142 | 0.2569 | 0.3046 | 99 |
| 1986 | 904610 | 321966 | 138659 | 70764 | 0.5103 | 0.5066 | 95 |
| 1987 | 2148730 | 392008 | 129575 | 44360 | 0.3423 | 0.3274 | 102 |
| 1988 | 927030 | 346130 | 155111 | 35591 | 0.2295 | 0.2729 | 97 |
| 1989 | 880110 | 331958 | 172862 | 34026 | 0.1968 | 0.2369 | 98 |
| 1990 | 438460 | 282168 | 164407 | 44693 | 0.2718 | 0.3316 | 101 |
| 1991 | 384680 | 217644 | 133997 | 28529 | 0.2129 | 0.2442 | 93 |
| 1992 | 799660 | 226087 | 110473 | 28985 | 0.2624 | 0.2706 | 99 |
| 1993 | 602380 | 192193 | 105026 | 31778 | 0.3026 | 0.2373 | 100 |
| 1994 | 855320 | 186474 | 96351 | 24430 | 0.2536 | 0.2216 | 100 |
| 1995 | 668080 | 165737 | 77034 | 29575 | 0.3839 | 0.2562 | 99 |
| 1996 | 866000 | 203681 | 123385 | 26105 | 0.2116 | 0.1668 | 95 |
| 1997 | 1538700 | 224251 | 79138 | 35233 | 0.4452 | 0.4801 | 99 |
| 1998 | 482370 | 189672 | 103762 | 33353 | 0.3214 | 0.4445 | 100 |
| 1999 | 307380 | 147028 | 87029 | 29736 | 0.3417 | 0.3089 | 99 |
| 2000 | 1726800 | 211001 | 75351 | 18322 | 0.2432 | 0.2065 | 100 |
| 2001 | 955630 | 224184 | 121998 | 24556 | 0.2013 | 0.2303 | 99 |
| 2002 | 921180 | 242966 | 137261 | 32914 | 0.2398 | 0.2476 | 99 |
| 2003 | 363590 | 198298 | 128104 | 28081 | 0.2192 | 0.2288 | 99 |
| 2004 | 259390 | 157963 | 110412 | 25021 | 0.2266 | 0.2002 | 98 |
| 2005 | 168210 | 118948 | 88261 | 14129 | 0.1601 | 0.1264 | 99 |
| 2006 | 112240 | 111054 | 77787 | 27346 | 0.3515 | 0.2755 | 100 |

No of years for separable analysis : 8
Age range in the analysis : 1 . . . 9
Year range in the analysis : 1958 . . . 2006
Number of indices of SSB : 0
Number of age-structured indices : 1
Parameters to estimate : 38
Number of observations : 216
Conventional single selection vector model to be fitted.

Table 5.6.15. Herring in VIa (N). Parameter estimates. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

PARAMETER ESTIMATES

| ${ }^{3}$ Parm. ${ }^{3}$ |  | 3 | Maximum | 3 | 3 | 3 | 3 | 3 | 3 | Mean of ${ }^{3}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 No. |  | 3 | Likelh. | 3 | CV | Lower | Upper | -s.e. | +s.e. | Param. |
| $3 \quad 3$ | 3 | 3 | Estimat | ${ }^{3}$ | (\%) ${ }^{3}$ | 95\% CL | 95\% CL | ${ }^{3}$ | 3 | Distrib. ${ }^{3}$ |
| Separable model : F by year |  |  |  |  |  |  |  |  |  |  |
| 1 | 1999 |  | 0.2669 |  | 16 | 0.1941 | 0.3669 | 0.2269 | 0.3139 | 0.2704 |
| 2 | 2000 |  | 0.1784 |  | 16 | 0.1294 | 0.2461 | 0.1514 | 0.2102 | 0.1808 |
| 3 | 2001 |  | 0.1989 |  | 16 | 0.1443 | 0.2742 | 0.1689 | 0.2343 | 0.2016 |
| 4 | 2002 |  | 0.2139 |  | 16 | 0.1537 | 0.2977 | 0.1807 | 0.2532 | 0.2170 |
| 5 | 2003 |  | 0.1977 |  | 17 | 0.1390 | 0.2811 | 0.1651 | 0.2366 | 0.2009 |
| 6 | 2004 |  | 0.1729 |  | 19 | 0.1179 | 0.2537 | 0.1422 | 0.2103 | 0.1763 |
| 7 | 2005 |  | 0.1092 |  | 21 | 0.0720 | 0.1655 | 0.0883 | 0.1350 | 0.1117 |
| 8 | 2006 |  | 0.2380 |  | 24 | 0.1466 | 0.3866 | 0.1859 | 0.3049 | 0.2454 |
| Separable Model: Selection (S) by age |  |  |  |  |  |  |  |  |  |  |
| 9 | 1 |  | 0.0087 |  | 38 | 0.0041 | 0.0186 | 0.0060 | 0.0128 | 0.0094 |
| 10 | 2 |  | 0.7272 |  | 15 | 0.5378 | 0.9833 | 0.6234 | 0.8482 | 0.7359 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| 12 | 5 |  | 1.3318 |  | 12 | 1.0369 | 1.7105 | 1.1722 | 1.5132 | 1.3427 |
| 13 | 6 |  | 1.0601 |  | 12 | 0.8320 | 1.3507 | 0.9368 | 1.1996 | 1.0682 |
| 14 | 7 |  | 1.0459 |  | 12 | 0.8178 | 1.3377 | 0.9226 | 1.1858 | 1.0542 |
| 81.0000 Fixed : Last true age |  |  |  |  |  |  |  |  |  |  |
| Separable model: Populations in year 2006 |  |  |  |  |  |  |  |  |  |  |
| 15 | 1 |  | 112248 |  | 94 | 17606 | 715632 | 43623 | 288831 | 175449 |
| 16 | 2 |  | 61823 |  | 34 | 31313 | 122063 | 43694 | 87475 | 65661 |
| 17 | 3 |  | 65199 |  | 27 | 37852 | 112302 | 49403 | 86045 | 67757 |
| 18 | 4 |  | 62387 |  | 24 | 38568 | 100917 | 48812 | 79737 | 64294 |
| 19 | 5 |  | 116367 |  | 22 | 74576 | 181577 | 92735 | 146022 | 119404 |
| 20 | 6 |  | 84945 |  | 22 | 54906 | 131419 | 67990 | 106128 | 87077 |
| 21 | 7 |  | 109862 |  | 21 | 71478 | 168857 | 88228 | 136800 | 112536 |
| 22 | 8 |  | 14511 |  | 21 | 9469 | 22238 | 11671 | 18043 | 14860 |
| Separable model: Populations at age |  |  |  |  |  |  |  |  |  |  |
| 23 | 1999 |  | 5959 |  | 29 | 3324 | 10681 | 4424 | 8026 | 6229 |
| 24 | 2000 |  | 10435 |  | 23 | 6523 | 16691 | 8211 | 13261 | 10739 |
| 25 | 2001 |  | 19125 |  | 21 | 12623 | 28975 | 15472 | 23640 | 19559 |
| 26 | 2002 |  | 18220 |  | 19 | 12332 | 26919 | 14930 | 22235 | 18585 |
| 27 | 2003 |  | 32094 |  | 20 | 21614 | 47657 | 26232 | 39267 | 32754 |
| 28 | 2004 |  | 58333 |  | 20 | 39170 | 86871 | 47607 | 71476 | 59550 |
| 29 | 2005 |  | 19283 |  | 20 | 12780 | 29094 | 15633 | 23785 | 19712 |

Age-structured index catchabilities
FLT01:West Scotland Summer Acoustic Sur

| Linear model | fitted. Slopes at age | : |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 30 | 1 | $Q$ | .6517 | 65 | .3483 | 4.498 | .6517 | 2.404 |
| 31 | 2 | $Q$ | 3.270 | 21 | 2.663 | 6.158 | 3.270 | 5.015 |

Table 5.6.16. Herring in VIa (N). Residuals about the model fit. N.B. In this table "age" refers to number of rings (winter rings in the otolith).



## Table 5.6.17. Herring in VIa (N). Parameters of distributions. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

PARAMETERS OF THE DISTRIBUTION OF $\ln (C A T C H E S ~ A T ~ A G E)$

| Separable model fitted from 1999 to 2006 |  |
| :--- | ---: |
| Variance | 0.2327 |
| Skewness test stat. | 0.1837 |
| Kurtosis test statistic | -0.1246 |
| Partial chi-square | 0.9624 |
| Significance in fit | 0.0000 |
| Degrees of freedom | 34 |

Degrees of freedom


PARAMETERS OF THE DISTRIBUTION OF THE AGE-STRUCTURED INDICES
DISTRIBUTION STATISTICS FOR FLT01:West Scotland Summer Acoustic Sur

| Linear catchability relationship | assumed | 2 |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1 | 2 | 4 | 5 | 6 | 7 | 8 |  |
| Variance | 0.0277 | 0.0324 | 0.0208 | 0.0286 | 0.0234 | 0.0393 | 0.0430 | 0.0404 |
| Skewness test stat. | -2.1467 | 2.8769 | 0.5223 | 0.5584 | -0.5677 | 0.2240 | -1.3030 | 0.8948 |
| Kurtosis test statisti | 1.2959 | 3.0106 | 0.4148 | -0.1743 | -0.2572 | -0.1417 | 0.6315 | -0.6440 |
| Partial chi-square | 0.0364 | 0.0414 | 0.0259 | 0.0361 | 0.0306 | 0.0512 | 0.0616 | 0.0596 |
| Significance in fit | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| Number of observations | 17 | 17 | 17 | 17 | 17 | 17 | 17 | 170.0000 |
| Degrees of freedom | 16 | 16 | 16 | 16 | 16 | 16 | 17 | 17 |
| Weight in the analysis | 0.0111 | 0.1111 | 0.1111 | 0.1111 | 0.1111 | 0.1111 | 0.1111 | 0.1111 |

## Table 5.6.18. Herring in VIa (N). Analysis of variance. N.B. In this table "age" refers to number of rings (winter rings in the otolith)

Unweighted Statistics
Variance
Total for model
Catches at age

| SSQ | Data | Parameters | d.f. | Variance |
| :---: | ---: | ---: | ---: | ---: |
| 114.2258 | 216 | 38 | 178 | 0.6417 |
| 22.9376 | 63 | 29 | 34 | 0.6746 |

Aged Indices
FLT01:West Scotland Summer Acoustic Su 91.2882
153
$\begin{array}{lll}9 & 144 & 0.6339\end{array}$
Weighted Statistics
Variance
Total for model
Catches at age

$$
\begin{array}{lr}
\text { SSQ } & \text { Data } \\
8.5512 & 216 \\
7.9126 & 63
\end{array}
$$

Parameters d.f. Variance

Aged Indices
FLT01:West Scotland Summer Acoustic Su 0.6387
153
$\begin{array}{lll}9 & 144 & 0.0044\end{array}$

Table 5.7.1.1. Herring in VIa (N). Input data for short-term predictions, numbers at age from the assessment with ages 1 and 2 replaced by geometric mean values - natural mortality (M), proportion mature (Mat), proportion of fishing mortality prior to spawning (PF), proportion of natural mortality prior to spawning (PM), mean weights at age in the stock (SWt), selection pattern (Sel), mean weights at age in the catch (CWt). All biological data are taken as mean of the last 3 years. VIa ( N ) herring appears to have considerable annual variability in mean weights and in fraction mature. The terminal year values are not applicable. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| 2007 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 1 | 604957 | 1 | 0 | 0.67 | 0.67 | 0.0697 | 2.08E-03 | 0.0664 |
| 2 | 222089 | 0.3 | 0.81 | 0.67 | 0.67 | 0.1342 | 0.173093 | 0.148233 |
| 3 | 38521 | 0.2 | 0.965 | 0.67 | 0.67 | 0.1596 | 0.29474 | 0.167933 |
| 4 | 39755 | 0.1 | 1 | 0.67 | 0.67 | 0.177167 | 0.238033 | 0.1908 |
| 5 | 44494 | 0.1 | 1 | 0.67 | 0.67 | 0.187067 | 0.317018 | 0.211467 |
| 6 | 76688 | 0.1 | 1 | 0.67 | 0.67 | 0.199033 | 0.252339 | 0.223633 |
| 7 | 59721 | 0.1 | 1 | 0.67 | 0.67 | 0.212167 | 0.248971 | 0.2258 |
| 8 | 77499 | 0.1 | 1 | 0.67 | 0.67 | 0.213667 | 0.238033 | 0.2357 |
| 9 | 28319 | 0.1 | 1 | 0.67 | 0.67 | 0.2218 | 0.238033 | 0.278533 |
| 2008 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 1 | 604957 | 1 | 0 | 0.67 | 0.67 | 0.0697 | 2.08E-03 | 0.0664 |
| 2 | . | 0.3 | 0.81 | 0.67 | 0.67 | 0.1342 | 0.173093 | 0.148233 |
| 3 | - | 0.2 | 0.965 | 0.67 | 0.67 | 0.1596 | 0.29474 | 0.167933 |
| 4 | - | 0.1 | 1 | 0.67 | 0.67 | 0.177167 | 0.238033 | 0.1908 |
| 5 | . | 0.1 | 1 | 0.67 | 0.67 | 0.187067 | 0.317018 | 0.211467 |
| 6 | . | 0.1 | 1 | 0.67 | 0.67 | 0.199033 | 0.252339 | 0.223633 |
| 7 | . | 0.1 | 1 | 0.67 | 0.67 | 0.212167 | 0.248971 | 0.2258 |
| 8 | . | 0.1 | 1 | 0.67 | 0.67 | 0.213667 | 0.238033 | 0.2357 |
| 9 | . | 0.1 | 1 | 0.67 | 0.67 | 0.2218 | 0.238033 | 0.278533 |


| 2009 |  |  |  | M | PF | PM | SWt | Sel |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | N | M | Mat | PFt |  |  |  |  |
| 1 | 604957 | 1 | 0 | 0.67 | 0.67 | 0.0697 | $2.08 \mathrm{E}-03$ | 0.0664 |
| 2 | $\cdot$ | 0.3 | 0.81 | 0.67 | 0.67 | 0.1342 | 0.173093 | 0.148233 |
| 3 | $\cdot$ | 0.2 | 0.965 | 0.67 | 0.67 | 0.1596 | 0.29474 | 0.167933 |
| 4 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.177167 | 0.238033 | 0.1908 |
| 5 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.187067 | 0.317018 | 0.211467 |
| 6 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.199033 | 0.252339 | 0.223633 |
| 7 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.212167 | 0.248971 | 0.2258 |
| 8 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.213667 | 0.238033 | 0.2357 |
| 9 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.2218 | 0.238033 | 0.278533 |

Table 5.7.1.2. Herring in VIa (N). Short-term prediction single option table, status quo F. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Year: | 2007 | F multiplier: | 1 | Fbar: | 0.276 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0021 | 796 | 53 | 604957 | 42166 | 0 | 0 | 0 | 0 |
| 2 | 0.1731 | 30628 | 4540 | 222089 | 29804 | 179892 | 24142 | 131025 | 17583 |
| 3 | 0.2947 | 8956 | 1504 | 38521 | 6148 | 37173 | 5933 | 26685 | 4259 |
| 4 | 0.238 | 8030 | 1532 | 39755 | 7043 | 39755 | 7043 | 31698 | 5616 |
| 5 | 0.317 | 11534 | 2439 | 44494 | 8323 | 44494 | 8323 | 33648 | 6294 |
| 6 | 0.2523 | 16310 | 3647 | 76688 | 15263 | 76688 | 15263 | 60563 | 12054 |
| 7 | 0.249 | 12552 | 2834 | 59721 | 12671 | 59721 | 12671 | 47270 | 10029 |
| 8 | 0.238 | 15653 | 3689 | 77499 | 16559 | 77499 | 16559 | 61793 | 13203 |
| 9 | 0.238 | 5720 | 1593 | 28319 | 6281 | 28319 | 6281 | 22580 | 5008 |
| Total |  | 110177 | 21832 | 1192043 | 144259 | 543541 | 96215 | 415260 | 74047 |
|  |  |  |  |  |  |  |  |  |  |
| Year: | 2008 | F multiplier: | 0.907 | Fbar: | 0.25 |  |  |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0019 | 722 | 48 | 604957 | 42166 | 0 | 0 | 0 | 0 |
| 2 | 0.157 | 27987 | 4149 | 222088 | 29804 | 179892 | 24141 | 132445 | 17774 |
| 3 | 0.2673 | 29551 | 4963 | 138377 | 22085 | 133534 | 21312 | 97636 | 15583 |
| 4 | 0.2159 | 4348 | 830 | 23487 | 4161 | 23487 | 4161 | 19007 | 3367 |
| 5 | 0.2875 | 6758 | 1429 | 28352 | 5304 | 28352 | 5304 | 21869 | 4091 |
| 6 | 0.2289 | 5719 | 1279 | 29322 | 5836 | 29322 | 5836 | 23523 | 4682 |
| 7 | 0.2258 | 10390 | 2346 | 53915 | 11439 | 53915 | 11439 | 43341 | 9196 |
| 8 | 0.2159 | 7799 | 1838 | 42128 | 9001 | 42128 | 9001 | 34092 | 7284 |
| 9 | 0.2159 | 13970 | 3891 | 75466 | 16738 | 75466 | 16738 | 61071 | 13546 |
| Total |  | 107244 | 20772 | 1218093 | 146534 | 566096 | 97933 | 432985 | 75523 |


| Year: | 2009 | F multiplier: | 0.907 | Fbar: | 0.25 |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0019 | 722 | 48 | 604957 | 42166 | 0 | 0 | 0 | 0 |
| 2 | 0.157 | 27992 | 4149 | 222131 | 29810 | 179926 | 24146 | 132471 | 17778 |
| 3 | 0.2673 | 30031 | 5043 | 140623 | 22443 | 135701 | 21658 | 99220 | 15836 |
| 4 | 0.2159 | 16053 | 3063 | 86718 | 15363 | 86718 | 15363 | 70176 | 12433 |
| 5 | 0.2875 | 4082 | 863 | 17125 | 3204 | 17125 | 3204 | 13209 | 2471 |
| 6 | 0.2289 | 3753 | 839 | 19243 | 3830 | 19243 | 3830 | 15438 | 3073 |
| 7 | 0.2258 | 4067 | 918 | 21104 | 4478 | 21104 | 4478 | 16965 | 3599 |
| 8 | 0.2159 | 7205 | 1698 | 38923 | 8317 | 38923 | 8317 | 31498 | 6730 |
| 9 | 0.2159 | 15873 | 4421 | 85742 | 19018 | 85742 | 19018 | 69387 | 15390 |
| Total |  | 109778 | 21044 | 1236567 | 148628 | 584483 | 100013 | 448364 | 77309 |

Table 5.7.1.3. Herring in VIa (N). Short-term prediction multiple option table,. status quo F.

| 2007 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings |  |  |
| 144259 | 74047 | 1 | 0.276 | 21832 |  |  |
| 2008 |  |  |  |  | 2009 |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 146534 | 87464 | 0 | 0 | 0 | 167097 | 106296 |
| . | 86186 | 0.0907 | 0.025 | 2286 | 165059 | 102908 |
| . | 84928 | 0.1814 | 0.05 | 4522 | 163066 | 99639 |
| . | 83689 | 0.2721 | 0.075 | 6711 | 161117 | 96486 |
| . | 82469 | 0.3628 | 0.1 | 8852 | 159212 | 93445 |
| . | 81267 | 0.4535 | 0.125 | 10948 | 157348 | 90510 |
| . | 80083 | 0.5442 | 0.1499 | 12999 | 155525 | 87678 |
| . | 78917 | 0.6349 | 0.1749 | 15006 | 153743 | 84946 |
| . | 77768 | 0.7256 | 0.1999 | 16969 | 152000 | 82309 |
| . | 76637 | 0.8163 | 0.2249 | 18891 | 150295 | 79765 |
| . | 75523 | 0.907 | 0.25 | 20772 | 148628 | 77309 |
| . | 74425 | 0.9977 | 0.2749 | 22613 | 146997 | 74939 |
| . | 73344 | 1.0884 | 0.2999 | 24415 | 145402 | 72652 |
| . | 72279 | 1.1791 | 0.3249 | 26178 | 143842 | 70444 |
| . | 71231 | 1.2698 | 0.3499 | 27904 | 142317 | 68312 |
| . | 70198 | 1.3605 | 0.3749 | 29594 | 140824 | 66254 |
| . | 69180 | 1.4512 | 0.3999 | 31247 | 139365 | 64267 |
| . | 68178 | 1.5419 | 0.4248 | 32866 | 137937 | 62349 |
| . | 67191 | 1.6326 | 0.4498 | 34450 | 136540 | 60497 |
| . | 66218 | 1.7233 | 0.4748 | 36001 | 135174 | 58708 |
| . | 65260 | 1.814 | 0.4998 | 37519 | 133838 | 56981 |

Table 5.7.1.4. Herring in VIa (N). Short-term prediction single option table, with TAC constraint. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Year: <br> Age | $\begin{array}{r} 2007 \\ \text { F } \end{array}$ | F multiplier: CatchNos | $\begin{array}{r} 1.6789 \\ \text { Yield } \end{array}$ | Fbar: <br> StockNos | $\begin{array}{r} 0.46 \\ \text { Biomass } \end{array}$ | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0035 | 1335 | 89 | 604957 | 42166 | 0 | 0 | 0 | 0 |
| 2 | 0.2906 | 48739 | 7225 | 222089 | 29804 | 179892 | 24142 | 121104 | 16252 |
| 3 | 0.4948 | 13740 | 2307 | 38521 | 6148 | 37173 | 5933 | 23337 | 3725 |
| 4 | 0.3996 | 12504 | 2386 | 39755 | 7043 | 39755 | 7043 | 28445 | 5040 |
| 5 | 0.5322 | 17552 | 3712 | 44494 | 8323 | 44494 | 8323 | 29130 | 5449 |
| 6 | 0.4236 | 25291 | 5656 | 76688 | 15263 | 76688 | 15263 | 53996 | 10747 |
| 7 | 0.418 | 19483 | 4399 | 59721 | 12671 | 59721 | 12671 | 42209 | 8955 |
| 8 | 0.3996 | 24376 | 5746 | 77499 | 16559 | 77499 | 16559 | 55452 | 11848 |
| 9 | 0.3996 | 8907 | 2481 | 28319 | 6281 | 28319 | 6281 | 20263 | 4494 |
| Total |  | 171928 | 34000 | 1192043 | 144259 | 543541 | 96215 | 373935 | 66510 |
| Year: | 2008 | F multiplier: | 0.907 | Fbar: | 0.25 |  |  |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0019 | 722 | 48 | 604957 | 42166 | 0 | 0 | 0 | 0 |
| 2 | 0.157 | 27947 | 4143 | 221775 | 29762 | 179637 | 24107 | 132258 | 17749 |
| 3 | 0.2673 | 26275 | 4412 | 123036 | 19637 | 118729 | 18949 | 86811 | 13855 |
| 4 | 0.2159 | 3559 | 679 | 19228 | 3407 | 19228 | 3407 | 15560 | 2757 |
| 5 | 0.2875 | 5750 | 1216 | 24122 | 4512 | 24122 | 4512 | 18606 | 3480 |
| 6 | 0.2289 | 4612 | 1031 | 23644 | 4706 | 23644 | 4706 | 18968 | 3775 |
| 7 | 0.2258 | 8755 | 1977 | 45427 | 9638 | 45427 | 9638 | 36518 | 7748 |
| 8 | 0.2159 | 6586 | 1552 | 35577 | 7602 | 35577 | 7602 | 28790 | 6152 |
| 9 | 0.2159 | 11886 | 3311 | 64206 | 14241 | 64206 | 14241 | 51958 | 11524 |
| Total |  | 96091 | 18369 | 1161970 | 135669 | 510570 | 87162 | 389470 | 67040 |
| Year: | 2009 | F multiplier: | 0.907 | Fbar: | 0.25 |  |  |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0019 | 722 | 48 | 604957 | 42166 | 0 | 0 | 0 | 0 |
| 2 | 0.157 | 27992 | 4149 | 222131 | 29810 | 179926 | 24146 | 132471 | 17778 |
| 3 | 0.2673 | 29988 | 5036 | 140424 | 22412 | 135509 | 21627 | 99080 | 15813 |
| 4 | 0.2159 | 14273 | 2723 | 77103 | 13660 | 77103 | 13660 | 62396 | 11054 |
| 5 | 0.2875 | 3342 | 707 | 14020 | 2623 | 14020 | 2623 | 10814 | 2023 |
| 6 | 0.2289 | 3193 | 714 | 16372 | 3259 | 16372 | 3259 | 13134 | 2614 |
| 7 | 0.2258 | 3280 | 741 | 17018 | 3611 | 17018 | 3611 | 13680 | 2902 |
| 8 | 0.2159 | 6071 | 1431 | 32795 | 7007 | 32795 | 7007 | 26539 | 5671 |
| 9 | 0.2159 | 13468 | 3751 | 72755 | 16137 | 72755 | 16137 | 58877 | 13059 |
| Total |  | 102330 | 19300 | 1197575 | 140683 | 545498 | 92070 | 416991 | 70914 |

Table 5.7.1.5. Herring in VIa (N). Short-term prediction multiple option table, with TAC constraint.

| 2007 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings |  |  |
| 144259 | 66510 | 1.6789 | 0.46 | 34000 |  |  |
| 2008 |  |  |  |  | 2009 |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 135669 | 77538 | 0 | 0 | 0 | 157046 | 96900 |
| - | 76415 | 0.0907 | 0.025 | 2020 | 155242 | 93866 |
| . | 75309 | 0.1814 | 0.05 | 3996 | 153477 | 90940 |
| . | 74220 | 0.2721 | 0.075 | 5931 | 151752 | 88115 |
| . | 73148 | 0.3628 | 0.1 | 7824 | 150064 | 85389 |
| . | 72091 | 0.4535 | 0.125 | 9677 | 148413 | 82759 |
| . | 71050 | 0.5442 | 0.1499 | 11491 | 146798 | 80220 |
| . | 70025 | 0.6349 | 0.1749 | 13266 | 145218 | 77769 |
| . | 69015 | 0.7256 | 0.1999 | 15003 | 143673 | 75403 |
| . | 68020 | 0.8163 | 0.2249 | 16704 | 142162 | 73119 |
| . | 67040 | 0.907 | 0.25 | 18369 | 140683 | 70914 |
| . | 66075 | 0.9977 | 0.2749 | 19999 | 139237 | 68785 |
| . | 65124 | 1.0884 | 0.2999 | 21594 | 137822 | 66730 |
| . | 64188 | 1.1791 | 0.3249 | 23156 | 136438 | 64745 |
| . | 63265 | 1.2698 | 0.3499 | 24685 | 135084 | 62828 |
| . | 62356 | 1.3605 | 0.3749 | 26182 | 133760 | 60977 |
| . | 61460 | 1.4512 | 0.3999 | 27647 | 132464 | 59189 |
| . | 60578 | 1.5419 | 0.4248 | 29082 | 131196 | 57462 |
| . | 59709 | 1.6326 | 0.4498 | 30486 | 129955 | 55794 |
| . | 58853 | 1.7233 | 0.4748 | 31861 | 128742 | 54182 |
| . | 58010 | 1.814 | 0.4998 | 33208 | 127554 | 52625 |



Figure 5.1. Herring in VIa (North). Map to show place names to the west of Scotland of relevance to the fishery and survey descriptions.


Figure 5.6.1. Herring in VIa (North). Herring in VIa (North). Separable model residual plots for the two exploratory assessments with data from 1958-2006. Left panels have the 1997 and 2005 acoustic surveys excluded; right panels include the 1997 and 2005 surveys.


Figure 5.6.2. Herring in VIa (North). Survey residual plots for the two exploratory assessments with data from 1958-2006. Left panels have the 1997 and 2005 acoustic surveys excluded; right panels include the 1997 and 2005 surveys.


Figure 5.6.3. Herring in VIa (North). Plot to show the value of reference $\mathbf{F}$ (and $\mathbf{9 5 \%}$ confidence intervals) obtained from the two exploratory assessment runs with the 1997 and 2005 acoustic surveys both excluded and included.



Figure 5.6.4. Herring in VIa (North). F and SSB from the two assessment runs both excluding and including the 1997 and 2005 acoustic surveys.




Figure 5.6.5. Herring in VIa (North). Analytical retrospective patterns (2006 to 2002) of SSB, mean F3-6 and recruitment from the assessments both excluding (dashed thicker line) and including (solid thicker line) the 1997 and 2005 acoustic surveys.


Figure 5.6.6. Herring in VIa (North). Illustration of stock trends from the assessment (8 year separable period). Summary of estimates of landings, fishing mortality at $\mathrm{F}_{3-6}$, recruitment at 1ring, spawning stock biomass at spawning time in the final assessment run including both the 1997 and 2005 acoustic survey.


Figure 5.6.7. Herring in VIa (North). Herring in VIa (N). Illustration of selection patterns diagnostics, from deterministic calculation (8-year separable period). Top left, a contour plot of selection pattern residuals. Top right, estimated selection (relative to 4 -ringers) +/- standard deviation. Bottom, marginal totals of residuals by year and ring.



MFYPR version 2 a
Run: testB Fstqu
Time and date: 11:16 20/03/2007

| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(3-6) | 1.0000 | 0.2755 |
| FMax | 65.0334 | 17.9188 |
| F0.1 | 0.6409 | 0.1766 |
| F35\%SPR | 0.6052 | 0.1667 |
| Flow | 0.3107 | 0.0856 |
| Fmed | 1.0094 | 0.2781 |
| Fhigh | 2.6750 | 0.7371 |

MFDP version 1a
Run: test3A
Herring Vla (north) (run: ICAPGF08/I08)
Time and date: 18:25 19/03/2007
Fbar age range: 3-6
Input units are thousands and kg - output in tonnes

Weights in kilograms

Figure 5.7.2.1. Herring in VIa (North). Yield-per-recruit and short-term forecast from the assessment including the 2005 acoustic survey.

## 6 Herring in Divisions VIa (South) and VIIb,c

### 6.1 The Fishery

### 6.1.1 Advice and management applicable to 2006-2007

The TAC for this area in 2006 was 15400 t with a decrease to 13860 t in 2007. For 2007, ICES advised fishing should not be allowed unless accompanied by a recovery plan, and stated that such a plan would include further reductions in catch.

In 2006, ICES considered the current SSB to be unknown, but likely to be below $\mathrm{B}_{\mathrm{pa}}$ (110 000 t ). For SSB to be above $\mathrm{B}_{\lim }(81000 \mathrm{t}$ ) there would have to have been very strong recruitment in recent years but no evidence has been found for such year classes. Fishing mortality has reduced in recent years but is likely to be above $\mathrm{F}_{\mathrm{pa}}(0.22)$ and $\mathrm{F}_{\text {lim. }}$ ( 0.33 ). This stock has been classified as being outside safe biological limits.

In 2000, an Irish committee was established to manage the Irish fishery for this stock. The committee stated its intention to follow the following objectives:

- To rebuild this stock to above the $\mathrm{B}_{\mathrm{pa}}$ level of 110000 t .
- In the event of the stock remaining below this level, additional conservation measures will need to be implemented.
- In the longer term it is the policy of the committee to further rebuild the stock to the level at which it can sustain annual catches of around 25000 t .
- Implement a closed season from March to October.
- Regulate effort further through boat quotas allocated on a weekly basis in the open season.

These objectives form the basis for the recovery plan that is being developed for this stock. STECF was asked to provide a preliminary evaluation of these objectives and stated that the aims outlined in this plan would be difficult to attain. Rebuilding the stock to above $\mathrm{B}_{\mathrm{pa}}$ would not be achievable given that there have been no strong year classes recruited to this fishery in recent years. Maintaining annual catches of 25000 t also would not be viable under current conditions. STECF concluded that no further TAC increases should be allowed if these aims are to be achieved.

The Pelagic RAC recommended that a management plan similar to that in VIa North be developed and the finalisation of this plan should await the recommendations of the WESTHER project.

### 6.1.2 Catches in 2006

The working group estimates of landings recorded by each country from this fishery from 1988 - 2006 are given in Table 6.1.2.1. Irish catch estimates for this WG have been based on the preliminary official reported data from the EU Logbook Scheme. The total official catch recorded from logbooks for 2006 was almost 15000 t , compared with about 13500 t in 2005. The working group catches in these areas from 1970 -2006 are shown in Figure 6.1.2.1.

In 2007 landings data were revised with respect to reallocation of catches between area VIaS and VIaN, in the years 2000-2005. Before 2000, a comprehensive reallocation was used. For 2000-2005, various procedures were used. These attempted to deal with the increasing Irish catches along the $56^{\circ}$ line and opportunistic Irish catches of herring in VIaN during the $4^{\text {th }}$ and $1^{\text {st }}$ quarter mackerel fishery. In some years some catches were reallocated, while in others no reallocations were made. In 2007, it was considered that the most correct procedure was that
used before 2000. Therefore a retrospective reallocation has been conducted for the years 2000-2005. It does not adequately consider the Irish herring catches in VIaN, nor does the reallocation consider fishing along the $56^{\circ}$ line. However, in the absence of better information on Irish directed herring fishing in VIaN, this procedure provides the best possible method.

There were no estimates of discards reported for 2006 and anecdotal reports from the industry are that discarding is not a major problem in this fishery at the present time.

### 6.1.3 The fishery in 2006

A total of 48 boats categorised as follows took part in the fishery in 2006:

- 22 Pelagic RSW boats from 27 m to 134 m
- 6 polyvalent RSW boats 24 m to 33 m
- 8 polyvalent tank hold boats 19 m to 26 m
- 12 polyvalent dry hold boats 14 m to 27 m

Polyvalent is a term used to define part of the Irish fleet licensed to catch pelagic and demersal fish.

In 2006 the majority of catches were reported from quarters 1 and 4 in VIaS with comparatively small catches reported from VIIb. In the first quarter the season opened on the $2^{\text {nd }}$ of January and closed on the $15^{\text {th }}$ February. Fishing reopened in the fourth quarter on the $1^{\text {st }}$ October and closed on the $7^{\text {th }}$ of December when the quota was exhausted. The distribution of catches in 2006 strongly reflects the size of the quota, the allocation system and the licensing conditions of the boats. Therefore as the quota was small and quickly caught effort was skewed towards those spawning grounds closest to the fleet base of Killybegs. Because pelagic RSW vessels are not allowed inside the Irish 12 mile limits, spawning grounds within territorial waters are comparatively lightly targeted. Within VIIb many of the herring spawning grounds are within the six mile limit and therefore accessible only to polyvalent vessels that receive a comparatively small cut of the quota. These factors raise the question of whether the fishery and sampling programme is representative of the stock. A map showing the spawning grounds of herring in this area and current perception of herring movements is shown in Figure 6.1.3.1.

### 6.2 Biological composition of the catch

### 6.2.1 Catch in numbers-at-age

Catch-at-age data for this fishery are available since 1970 and are shown in Table 6.2.1.1 with percentages since 1994 shown in Table 6.2.1.2. One ringers are never well represented in the catch. Generally it is found that 1 ringers do not show up in the catch until quarter 4 . The proportions of 2, 3, and 4 ringers in 2006 accounted for $18 \%$, $29 \%$ and $25 \%$ respectively. 2006 shows a similar age profile to 2005 with a peak in 3 ringers.

Two ringers dominate the catch in quarter 4 while in quarter 1 , dominance is shared between 3 and 4 ringers. Overall quarter 1 had a greater proportion of older age classes which represent the larger spring spawners. There is little evidence for 1 ringers being an important component of landings in this area.

The length distributions of the catches taken per quarter by the Irish fleet are shown in Table 6.2.2.2. A particular aspect of the first quarter fishery in VIaS has been the appearance of large spring spawning fish off the north coast in late January and early February. These fish are usually over 31 cm in total length. Herring from quarter 1 show a wider length distribution $(20 \mathrm{~cm}-35.5 \mathrm{~cm})$ than those from quarter $4(21.5 \mathrm{~cm}-29.5 \mathrm{~cm})$.

### 6.2.2 Quality of the catch and biological data

The quality of landings data has improved in recent years, particularly since September 2004. At that time a new $2 \%$ tolerance for water in catches was also introduced. Before that time, the tolerance was $20 \%$. This suggests that the data pre and post 2004 are not directly comparable. Enforcement of the TAC improved since the late 1990s and again in 2004 and is now extremely strong.

Discarding, throughout the time series is not considered to be a big a problem, however estimates are not available. The fishery on this stock was not so reliant on roe and thus the incentive to discard was less than in the Celtic Sea fishery.

The numbers of samples and the associated biological data are shown in Table 6.2.2.1. Along with the need to obtain more samples from the larger RSW vessels there is also a requirement to sample the opportunistic catches that occur in VIIb.

Precision estimates for the Irish catch at age data for this stock are presented in Section 1.5.3. They show excellent levels of precision ( $\mathrm{CV}<4 \%$ ) for the main ages in the catches.

### 6.3 Fishery Independent Information

### 6.3.1 Ground Fish Surveys

There are currently no recruitment indices available for this stock. The western IBTS fourth quarter surveys (Ireland and UK-Scotland) offer the possibility of recruit indices. The Irish survey only began in 2003 and the time series is too short. These surveys should be investigated in future years, when the time series is longer. However the WESTHER project (2007) shows that juveniles from this stock and from VIaN mix together in areas such as Stanton Bank. This report also shows that juveniles of VIaS origin are found in Sea Lochs in VIaN. Therefore, survey information from VIaN, particularly the Scottish IBTS survey will be useful for this stock, if it is possible to identify the proportion of VIaS fish.

### 6.3.2 Acoustic Surveys

Acoustic surveys have been conducted in this area since 1994. In the mid 1990s, surveys were undertaken in summer. The timing changed in 1999 with the surveys being carried out in the winter. The 2007 survey was the $9^{\text {th }}$ in the current time series. The new series of winter surveys aims to measure the abundance of spawning and pre-spawning spring spawning components. This component is considered to be the dominant component at present. A problem with the winter acoustic survey series has been synchronising the survey with the peak spawning event to ensure maximal containment of the stock.

The January 2007 survey track and SA values attributed to herring are shown in Figures 6.3.2.1 and 6.3.2.2. Details of the acoustic surveys in this area are presented in Table 6.3.2.1. The survey started in the south, covering the main bays and inlets into the north of VIIb, and moved in a northerly direction to correspond with the timing of spawning, ending at Malin Head in VIaS. Poor weather affected the survey and significant alterations had to be made to the original cruise track.

The majority of fish recorded during the survey were mature, accounting for $98 \%$ of the biomass and $97 \%$ of the abundance. Spent fish accounted for $51 \%$ of the total biomass and total abundance. This dominance of spent fish would seem to indicate that the main spawning event in this region had already taken place and/or the spring spawning fish had not moved inshore to spawn yet. Less than $2 \%$ of the total stock biomass was made up of juvenile herring and less than $3 \%$ of the total stock numbers.

The age profile from this survey is similar to the 2006 survey with 3 and 4 ringers being the dominant year classes representing $56.7 \%$ and $20.8 \%$ by numbers respectively of the total stock biomass. Overall the numbers of older fish present in the biological samples has decreased. The age range present in the 2007 survey was 1-7 while in 2006 it was 1-9.

The age distribution of the abundance estimate from the acoustic survey and from the commercial fishery in 2006 is presented in Figure 6.3.2.3. The quarter one fishery shows slightly greater amounts of 4 ringers, with 3 and 5 ringers also present in significant amounts. The age profiles from the last four acoustic surveys are plotted in Figure 6.3.2.3.

The total biomass estimate for the area surveyed in 2007 was 14222 t with an SSB of 13974 t . This is a significant decrease from the 2006 estimate which was 27750 t with an SSB of 27200 t . Poor weather may have effected the quality of the estimate in 2007.

This survey is not considered a good indicator of stock development. This is because of the protracted nature of the spawning period, from October to at least February, whilst the survey is conducted in early January. Also it has been shown by the WESTHER project (Hatfield et al. 2007 WD) that some fish of VIaS origin are present in VIaN. These fish are picked up in the VIaN acoustic survey and are not accounted for in tuning the VIaS assessment. In light of the WESTHER results HAWG recommends that the utility of the VIaS winter survey be evaluated. A likely alternative is to conduct a survey in July corresponding with the VIaN survey, and extending south from $56^{\circ} 30 \mathrm{~N}$ to at least VIIb. Such a survey would be coordinated by PGHERS.

### 6.4 Mean weights-at-age and maturity-at-age

The mean weights $(\mathrm{kg})$ at age in the catches in 2006 are based on Irish catches and are very similar to 2005 for ringers 1-7 (Table 6.4.1.1). These mean weights display quite a stable pattern over the time series, although variable weights are only available from the early 1980s (Figure 6.4.4.1.).

The mean weights in the stock at spawning time have been calculated from Irish samples taken during the main spawning period that extends from October to February (Table 6.4.1.2). The time series since 1986 is shown in Figure 6.4.4.2.

A maturity ogive has been produced from the 2007 acoustic survey shows that $58 \%$ are mature at 1 -ring, $99 \%$ at 2 -ring and $100 \%$ mature at 3 -ring. The maturity ogive used in the assessment considers 1 -ringers to be all immature and all subsequent age groups as fully mature.

### 6.5 Recruitment

Recruits (1-ringers) are poorly represented in the catch at age data. In addition, they rarely appear in the fishery until quarter 4 . Thus there is little fishery dependent information on recruitment strength. The converged separable VPAs (Figure 6.6.2.1) show episodic good recruitments roughly every 3-4 years. Good recruitments often occurred in subsequent years. The 1981 and 1985 year classes were abnormally good, and the 1983 year class was also well above the strength of any others.

### 6.6 Stock Assessment

### 6.6.1 Trends and patterns in basic data

The numbers at age from the catch and the survey have been mean standardised by year and are presented in Figure 6.6.1.1. The timing of the acoustic surveys changed in 2003. The early
surveys were carried out in quarter 4 while the later ones were in quarter 1 . The survey age profile reflects this change in timing.

Since the mid nineties there has been a pronounced shift in the age composition of the catches from old fish to younger fish and this has been maintained in 2007, with $2-5$ ringers making up the bulk of the catch. The numbers of 1 ringers in the survey and the catch is very low in 2006 and 2007. As in previous years the catch numbers at age do not suggest strong incoming recruitment and there is no information in the survey abundance to refute this. The catch and survey data also show a decrease in the numbers of older fish.

The $\log$ catch ratios ( $\ln \mathrm{C}_{\mathrm{a}, \mathrm{y}} / \mathrm{C}_{\mathrm{a}+1, \mathrm{y}+1}$ ) are presented in Figure 6.6.1.2 and are smoothed with a 4 -year running average to show the main trends. Data for 1 -ringers are noisy because this group is not fully selected by the fishery. The data for older fish are also noisy, particularly in later years, reflecting their relative paucity in the catches and suggest high variability in the exploitation rates of these age groups. These show an upward trend for all fully recruited year classes since the mid nineties. Overall, the catch data show a diminishing range of ages in the catches and older fish are at their lowest levels in the time series.

Cohort catch curves, were constructed for each year class in the catch at age data (Figure 6.6.1.3). These catch curves show signals in total mortality over the time series. Low mortality seems evident on the very large 1981, 1985 and 1988 year classes. These represent three of the biggest year classes recruited to this fishery. Increasing mortality can be seen from 1990 on, whilst the 1970 s cohorts show lower Z .

Figure 6.6.1.4 shows the catch curves from fishery and the survey age data averaged over a number of years. Total mortality has displayed an upward trend since the early 1980s. Increasing mortality in recent years is clearly evident. In the late seventies mortality was at its lowest (0.2). Since the mid nineties mortality has shown an increase to 0.7 . The survey displays broadly similar mortality signal to the catch.

### 6.6.2 Exploratory Assessments

Following the procedure of recent years, a separable VPA was used to screen over three terminal fishing mortalities, $0.2,0.4$ and 0.6 . This was achieved using the Lowestoft VPA software (Darby and Flatman, 1994). Reference age for calculation of fishing mortality was 3-6 and terminal selection was fixed at 1 , relative to age 4 (winter rings). ICA was also used in 2007 with the split tuning series. This assessment is still exploratory, and no assessment has been accepted by ACFM in recent years.

Three assessments using the separable VPA are presented, based on the three choices of terminal F. Recruitment, SSB and mean F are plotted in Figure 6.6.2.1, with last year's results, for comparative purposes. This figure is more informative for the converged part of the VPA, but in most recent years has little information on the current stock dynamics. Outputs from separable VPAs with terminal Fs of 0.2, 0.4 and 0.6 are presented in Tables 6.6.2.1, 6.6.2.2 and 6.6.2.3 respectively. Residual plots for the three trial assessments are presented in Figure 6.6.2.2. A strong negative residual pattern can be seen in 6 ringers.

Under each scenario of terminal F, the current assessment suggests declining fishing mortality since 2002, with a sharp increase in 2006. The landings have been fluctuating in recent years with 2006 the highest since 2000.

Recruitment appears to have remained stable at a low level when $\mathrm{F}=0.4$ or 0.6 . A higher level of recruitment can be seen when except when $\mathrm{F}=0.2$. Each scenario shows recruitment to be at a similar level in the final year and this is calculated using the geometric mean of the recruitment index over the entire time series. These explorations are only useful as indicators of historic trends. These results are consistent with the preliminary data screening that shows
no stronger year classes in the fishery in recent years. However these 1-ringer fish are poorly selected in the fishery and thus there is little information in the catch at age matrix on their strength in the final year.

SSB is either stable at a low level or declining slightly, assuming terminal F of 0.4 or 0.6 and possibly increasing at F values of 0.2 . If SSB is stable, it is stable at the lowest level in the series and is considerably lower than the current levels of $\mathrm{B}_{\mathrm{pa}}$ and $\mathrm{B}_{\mathrm{lim}}$. If using an $\mathrm{F}=0.2$ is more realistic then SSB is at a level just at $\mathrm{B}_{\mathrm{lim}}$. Only a terminal F of 0.2 suggests some good recruitments in recent years however there is no evidence in the raw catch numbers at age to suggest that this is so.

A number of assessments were conducted in ICA using the same procedure as last year with the split survey time series using ages three and four. Results of recruitment, SSB and mean F from the ICA runs, compared with the VPA runs are presented in Figures 6.6.2.3. The 2007 acoustic survey is included in one run and excluded in the other. The ICA runs show similar patterns of recruitment and SSB. Inclusion of the 2007 survey shows mean $F$ at a higher level.

The separable model and survey residuals from the ICA runs are presented in Figure 6.6.2.4. The year and age residuals are presented in Figure 6.6.2.5. The magnitude and location of residuals shown in the bubble plots are consistent. A switch from positive to negative residuals can be seen in the age residuals when the 2007 survey is removed.

A scatter plot was produced, plotting F and SSB in the terminal year to present the precision of the different runs, with and without the 2007 survey (Figure 6.6.2.6). Inclusion of the 2007 survey shows more variation in F while exclusion of the survey shows more variation in SSB. HAWG therefore concluded that ICA was not an informative model with which to provide management advice.

### 6.7 Short term projections

In the absence of an agreed assessment, it was not considered informative to carry out any predictions.

### 6.8 Medium term projections

Yield per recruit analyses were performed in 2006, and it is not considered necessary to update them.

### 6.9 Precautionary and yield based reference points

In 2007 the technical basis for the selection of the precautionary reference points was examined based on methods used by SGPRP (ICES 2001 ACFM:11). No alternative biomass and fishing mortality reference points are available. It is clear that recruitment does not show any clear dependence on the SSB and that apart from the very high year classes in the 1980s is showing a decline. The SGPRP (ICES 2003/ACFM:15) has reviewed the methodology for the calculation of biological reference points, and applying a segmented regression to the stock and recruit data from the 2002 HAWG assessment showed that the fit to the stock and recruit data for this stock was not significant.

### 6.10 Quality of the Assessment

In light of current uncertainties no assessment was conducted.

### 6.11 Management Considerations

The results of the non-tuned assessment suggest that SSB may be stable at a low level but estimates of SSB for this stock are uncertain. Though the peak in SSB in the 1980s may have been an isolated event the HAWG suggests that this stock should be exploited with great caution.

Little information on recruitment is currently available and it is unlikely that it is above average. There is no evidence that large year classes have recruited to the stock in recent years. F appears to have increased concomitantly with increases in the catch. The TAC increased from 14000 t to 15400 t in 2006. The management of the Irish fishery (which takes most of the catch) has improved in recent years with stricter quota enforcement in place. Certainly every effort should be taken to maintain catches below the current level. HAWG notes that increased accuracy in the catch data over the past 4 years gives a greater confidence in the perception of stock development.

Management objectives were reviewed by STECF and HAWG agrees with this review. It was concluded that rebuilding the stock to levels above $\mathrm{B}_{\mathrm{pa}}$ and maintaining annual catches of $25,000 \mathrm{t}$ would not be achievable. HAWG commends aspects of the plan such as the regulation of effort through tight enforcement of catch quotas, and this should be continued and if necessary intensified. The closed season from March to October has been maintained and is also commended.

Table 6.1.2.1. VIa(S) and VIIb,c herring. Estimated Herring catches in tonnes, 1988-2006. These data do not in all cases correspond to the official statistics and cannot be used for management purposes.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | - | - | + | - | - | - | - | - | - | - |
| Germany, Fed.Rep. | - | - | - | - | 250 | - | - | 11 | - | - |
| Ireland | 15000 | 18200 | 25000 | 22500 | 26000 | 27600 | 24400 | 25450 | 23800 | 24400 |
| Netherlands | 300 | 2900 | 2533 | 600 | 900 | 2500 | 2500 | 1207 | 1800 | 3400 |
| UK (N.Ireland) | - | - | 80 | - | - | - | - | - | - |  |
| UK (England + Wales) | - | - | - | - | - | - | 50 | 24 | - | - |
| UK Scotland | - | + | - | + | - | 200 | - | - | - | - |
| Total landings | 15300 | 21100 | 27613 | 23100 | 27150 | 30300 | 26950 | 26692 | 25600 | 27800 |
| Unallocated/ area misreported | 13800 | 7100 | 13826 | 11200 | 4600 | 6250 | 6250 | 1100 | 6900 | -700 |
| Discards | - | 1000 | 2530 | 3400 | 100 | 250 | 700 | - | - | 50 |
| WG catch | 29100 | 29200 | 43969 | 37700 | 31850 | 36800 | 33900 | 27792 | 32500 | 27150 |
| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |  |
| France | - | - | - | - | 515 | - | - | - | - |  |
| Germany, Fed.Rep. | - | - | - | - | - | - | - | - | - |  |
| Ireland | 25200 | 16325 | 10164 | 11278 | 13072 | 12921 | 10950 | 13351 | 14840 |  |
| Netherlands | 2500 | 1868 | 1234 | 2088 | 366 | - | 64 | - | 353 |  |
| UK (N.Ireland) | - | - | - | - | - | - | - | - | - |  |
| UK (England + Wales) | - | - | - | - | - | - | - | - | - |  |
| UK Scotland | - | - | - | - | - | - | - | - | 6 |  |
| Total landings | 27700 | 18193 | 11398 | 13366 | 13953 | 12921 | 11014 | 13351 | 15199 |  |
| Area misreported/Unallocated | 11200 | 7916 | 8448 | 1390 | 3873 | 3581 | 2813 | 2880 | 3994 |  |
| Discards | - | - | - | - | - | - | - | - |  |  |
| WG catch | 38900 | 26109 | 19846 | 14756 | 17826 | 16502 | 13827 | 16231 | 19193 |  |

Table 6.2.1.1 VIa(S) \& VIIb,c herring. Catch in numbers-at-age (winter rings) from 1970 to 2006.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 135 | 35114 | 26007 | 13243 | 3895 | 40181 | 2982 | 1667 | 1911 |
| 1971 | 883 | 6177 | 7038 | 10856 | 8826 | 3938 | 40553 | 2286 | 2160 |
| 1972 | 1001 | 28786 | 20534 | 6191 | 11145 | 10057 | 4243 | 47182 | 4305 |
| 1973 | 6423 | 40390 | 47389 | 16863 | 7432 | 12383 | 9191 | 1969 | 50980 |
| 1974 | 3374 | 29406 | 41116 | 44579 | 17857 | 8882 | 10901 | 10272 | 30549 |
| 1975 | 7360 | 41308 | 25117 | 29192 | 23718 | 10703 | 5909 | 9378 | 32029 |
| 1976 | 16613 | 29011 | 37512 | 26544 | 25317 | 15000 | 5208 | 3596 | 15703 |
| 1977 | 4485 | 44512 | 13396 | 17176 | 12209 | 9924 | 5534 | 1360 | 4150 |
| 1978 | 10170 | 40320 | 27079 | 13308 | 10685 | 5356 | 4270 | 3638 | 3324 |
| 1979 | 5919 | 50071 | 19161 | 19969 | 9349 | 8422 | 5443 | 4423 | 4090 |
| 1980 | 2856 | 40058 | 64946 | 25140 | 22126 | 7748 | 6946 | 4344 | 5334 |
| 1981 | 1620 | 22265 | 41794 | 31460 | 12812 | 12746 | 3461 | 2735 | 5220 |
| 1982 | 748 | 18136 | 17004 | 28220 | 18280 | 8121 | 4089 | 3249 | 2875 |
| 1983 | 1517 | 43688 | 49534 | 25316 | 31782 | 18320 | 6695 | 3329 | 4251 |
| 1984 | 2794 | 81481 | 28660 | 17854 | 7190 | 12836 | 5974 | 2008 | 4020 |
| 1985 | 9606 | 15143 | 67355 | 12756 | 11241 | 7638 | 9185 | 7587 | 2168 |
| 1986 | 918 | 27110 | 24818 | 66383 | 14644 | 7988 | 5696 | 5422 | 2127 |
| 1987 | 12149 | 44160 | 80213 | 41504 | 99222 | 15226 | 12639 | 6082 | 10187 |
| 1988 | 0 | 29135 | 46300 | 41008 | 23381 | 45692 | 6946 | 2482 | 1964 |
| 1989 | 2241 | 6919 | 78842 | 26149 | 21481 | 15008 | 24917 | 4213 | 3036 |
| 1990 | 878 | 24977 | 19500 | 151978 | 24362 | 20164 | 16314 | 8184 | 1130 |
| 1991 | 675 | 34437 | 27810 | 12420 | 100444 | 17921 | 14865 | 11311 | 7660 |
| 1992 | 2592 | 15519 | 42532 | 26839 | 12565 | 73307 | 8535 | 8203 | 6286 |
| 1993 | 191 | 20562 | 22666 | 41967 | 23379 | 13547 | 67265 | 7671 | 6013 |
| 1994 | 11709 | 56156 | 31225 | 16877 | 21772 | 13644 | 8597 | 31729 | 10093 |
| 1995 | 284 | 34471 | 35414 | 18617 | 19133 | 16081 | 5749 | 8585 | 14215 |
| 1996 | 4776 | 24424 | 69307 | 31128 | 9842 | 15314 | 8158 | 12463 | 6472 |
| 1997 | 7458 | 56329 | 25946 | 38742 | 14583 | 5977 | 8351 | 3418 | 4264 |
| 1998 | 7437 | 72777 | 80612 | 38326 | 30165 | 9138 | 5282 | 3434 | 2942 |
| 1999 | 2392 | 51254 | 61329 | 34901 | 10092 | 5887 | 1880 | 1086 | 949 |
| 2000 | 4101 | 34564 | 38925 | 30706 | 13345 | 2735 | 1464 | 690 | 1602 |
| 2001 | 2316 | 21717 | 21780 | 17533 | 18450 | 9953 | 1741 | 1027 | 508 |
| 2002 | 4058 | 32640 | 37749 | 18882 | 11623 | 10215 | 2747 | 1605 | 644 |
| 2003 | 1731 | 32819 | 28714 | 24189 | 9432 | 5176 | 2525 | 923 | 303 |
| 2004 | 1401 | 15122 | 32992 | 19720 | 9006 | 4924 | 1547 | 975 | 323 |
| 2005 | 209 | 28123 | 30896 | 26887 | 10774 | 5452 | 1348 | 858 | 243 |
| 2006 | 598 | 22036 | 36700 | 30581 | 21956 | 9080 | 2418 | 832 | 369 |

Table 6.2.1.2 VIa(S) \& VIIb,c herring. Percentage age composition (winter rings).

|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ |
| :--- | :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 4}$ | 6 | 28 | 15 | 8 | 11 | 7 | 4 | 16 | 5 |
| $\mathbf{1 9 9 5}$ | 0 | 23 | 23 | 12 | 13 | 11 | 4 | 6 | 9 |
| $\mathbf{1 9 9 6}$ | 3 | 13 | 38 | 17 | 5 | 8 | 4 | 7 | 4 |
| $\mathbf{1 9 9 7}$ | 5 | 34 | 16 | 23 | 9 | 4 | 5 | 2 | 3 |
| $\mathbf{1 9 9 8}$ | 3 | 29 | 32 | 15 | 12 | 4 | 2 | 1 | 1 |
| $\mathbf{1 9 9 9}$ | 1 | 30 | 36 | 21 | 6 | 3 | 1 | 1 | 1 |
| $\mathbf{2 0 0 0}$ | 3 | 27 | 30 | 24 | 10 | 2 | 1 | 1 | 1 |
| $\mathbf{2 0 0 1}$ | 2 | 23 | 23 | 18 | 19 | 10 | 2 | 1 | 1 |
| $\mathbf{2 0 0 2}$ | 3 | 27 | 31 | 16 | 10 | 9 | 2 | 1 | 1 |
| $\mathbf{2 0 0 3}$ | 2 | 31 | 27 | 23 | 9 | 5 | 2 | 1 | 0 |
| $\mathbf{2 0 0 4}$ | 2 | 18 | 38 | 23 | 10 | 6 | 2 | 1 | 0 |
| $\mathbf{2 0 0 5}$ | 0 | 27 | 29 | 26 | 10 | 5 | 1 | 1 | 0 |
| $\mathbf{2 0 0 6}$ | 0 | 18 | 29 | 25 | 18 | 7 | 2 | 1 | 0 |

Table 6.2.2.1 VIa(S) and VIIb,c herring. Sampling intensity of catches in 2006.

| ICES area | Year | Quarter | Landings (t) | No. Samples | No. aged | No. Measured | Aged/1000 t |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| VIaS | 2006 | 1 | 7789 | 22 | 1400 | 4246 | 180 |
| VIaS | 2006 | 4 | 5419 | 25 | 1342 | 4706 | 248 |
| VIIb | 2006 | 1 | 675 | 5 | 286 | 1031 | 424 |
| VIIb | 2006 | 4 | 957 | 2 | 101 | 260 | 106 |
|  |  |  |  |  |  |  |  |
| Total North West |  | 14840 | 54 | 3129 | 10243 | 957 |  |

Table 6.2.2.2. VIa(S) and VIIb,c herring. Length distribution of Irish catches/quarter (thousands) 2006.

| Length cm | Quarter 1 | Quarter 1 | Quarter 4 | Quarter 4 |
| :---: | :---: | :---: | :---: | :---: |
|  | VIaS | VIII, c | VIaS | VIIb,c |
| 20 | 39 |  |  |  |
| 20.5 | 39 |  |  |  |
| 21 | 91 |  |  |  |
| 21.5 | 220 | 48 | 4 |  |
| 22 | 220 | 132 | 4 |  |
| 22.5 | 414 | 346 | 4 | 43 |
| 23 | 518 | 575 | 22 | 43 |
| 23.5 | 1269 | 803 | 138 | 86 |
| 24 | 2111 | 1642 | 299 | 107 |
| 24.5 | 2435 | 1981 | 464 | 236 |
| 25 | 3678 | 2757 | 477 | 364 |
| 25.5 | 4222 | 3061 | 491 | 450 |
| 26 | 5063 | 3470 | 584 | 514 |
| 26.5 | 6721 | 4745 | 620 | 599 |
| 27 | 7822 | 4675 | 696 | 921 |
| 27.5 | 7200 | 3595 | 446 | 856 |
| 28 | 5789 | 2590 | 201 | 792 |
| 28.5 | 3302 | 1281 | 98 | 321 |
| 29 | 1619 | 616 | 40 | 171 |
| 29.5 | 777 | 208 | 9 | 43 |
| 30 | 453 | 55 |  | 21 |
| 30.5 | 207 | 14 |  |  |
| 31 | 233 | 0 |  |  |
| 31.5 | 78 |  |  |  |
| 32 | 78 |  |  |  |
| 32.5 | 78 |  |  |  |
| 33 | 52 |  |  |  |
| 33.5 | 91 |  |  |  |
| 34 | 78 |  |  |  |
| 34.5 | 39 |  |  |  |
| 35 | 13 |  |  |  |
| 35.5 | 39 |  |  |  |
| Nos./t | 7059 | 6015 | 6811 | 5817 |

Table 6.3.2.1. VIa(S) \& VIIb,c herring. Details of acoustic surveys of herring in VIaS and VIIbc, 1994-2007.

| Year | Type | Biomass | SSB | Reference |
| :--- | :--- | :---: | :---: | :--- |
|  |  |  |  |  |
| $\mathbf{1 9 9 4}$ | Feeding phase | - | 353,772 | Fernandes, 1994 |
| $\mathbf{1 9 9 5}$ | Feeding phase | 137,670 | 125,800 | Fernandes, 1995 |
| $\mathbf{1 9 9 6}$ | Feeding phase | 34,290 | 12,550 | Fernandes, 1996 |
| $\mathbf{1 9 9 7}$ | - | - | - | - |
| $\mathbf{1 9 9 8}$ | - | - | - | - |
| $\mathbf{1 9 9 9}$ | Autumn spawners | 23,762 | 22,788 | Breslin, 1999 |
| $\mathbf{2 0 0 0}$ | Autumn spawners | 21,000 | 20,500 | Breslin and Griffin, 2001 |
| $\mathbf{2 0 0 1}$ | Autumn spawners | 11,100 | 9,800 | Breslin and Griffin, 2002 |
| $\mathbf{2 0 0 2}$ | Winter spawners | 8,900 | 7,200 | Breslin and Griffin, 2003 |
| $\mathbf{2 0 0 3}$ | Winter spawners | 10,300 | 9,500 | Breslin and Griffin, 2003 |
| $\mathbf{2 0 0 4}$ | Winter spawners | 41,700 | 41,399 | Griffin, 2004 |
| $\mathbf{2 0 0 5}$ | Winter spawners | 71,253 | 66,138 | O Donnell et al., 2005 |
| $\mathbf{2 0 0 6}$ | Winter spawners | 27,770 | 27,200 | O Donnell et al., 2006 |
| $\mathbf{2 0 0 7}$ | Winter spawners | 14,222 | 13,974 | O Donnell et al., 2007 |

Table 6.3.2.2. VIa(S) \& VIIb,c herring. Time series of acoustic surveys since 1999.

| Winter rings | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |
| $\mathbf{0}$ | - | - | 5 | 0 | - | 0.09 | 1.28 | 0 | - |
| $\mathbf{1}$ | 18.99 | 10.71 | 22.69 | 35.7 | 10.28 |  | 7.83 | 1.6 | 0.3 |
| $\mathbf{2}$ | 104.77 | 60.88 | 52.33 | 14.05 | 26.26 | 3.9 | 56.91 | 6.9 | 3.5 |
| $\mathbf{3}$ | 32.53 | 48.96 | 6.41 | 24.23 | 30.02 | 62.35 | 93.51 | 86.7 | 59.8 |
| $\mathbf{4}$ | 11.34 | 25.57 | 6.47 | 14 | 11.08 | 54.93 | 109.87 | 57.5 | 21.9 |
| $\mathbf{5}$ | 1.65 | 9.43 | 2.63 | 5.79 | 2.94 | 80.07 | 100.8 | 27.9 | 11.7 |
| $\mathbf{6}$ | 0.94 | 2.35 | 1.94 | 5.7 | 0.64 | 47.14 | 56.54 | 16 | 6.35 |
| $\mathbf{7}$ | 0.3 | 1.28 | 0.12 | 5.06 | 0.94 | 13.81 | 21.16 | 4.8 | 1.86 |
| $\mathbf{8}$ | 0.17 | 0.43 | 0.24 | 2.73 | 0.3 | 11.77 | 24.64 | 4.8 | - |
| 9+ | 0.11 | 0.75 | 0.07 | 4.07 | 0.14 | - | 12.74 | 1.3 | - |
|  |  |  |  |  |  |  |  |  |  |
| Abundance (millions) | 170.8 | 160.36 | 97.9 | 111.33 | 82.6 | 274.06 | 485.29 | 202.9 | 105.41 |
| Total Biomass (t) | 23,762 | 21,048 | 11,062 | 8,867 | 10,300 | 41,700 | 71,253 | 27,770 | 14,222 |
| SSB (t) | 22,788 | 20,500 | 9,800 | 6,978 | 9,500 | 41,300 | 66,138 | 27,200 | 13,974 |
| CV | - | - | - | - | - | - | - | $49 \%$ | $44 \%$ |

Table 6.4.1.1. VIa(S) \& VIIb,c herring. Mean weight-at-age (winter rings) in the catch, 1970 to 2006.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1971 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1972 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1973 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1974 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1975 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1976 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1977 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1978 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1979 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1980 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1981 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1982 | 0.110 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1983 | 0.090 | 0.129 | 0.165 | 0.191 | 0.209 | 0.222 | 0.231 | 0.237 | 0.241 |
| 1984 | 0.106 | 0.141 | 0.181 | 0.210 | 0.226 | 0.237 | 0.243 | 0.247 | 0.248 |
| 1985 | 0.077 | 0.122 | 0.161 | 0.184 | 0.196 | 0.206 | 0.212 | 0.225 | 0.230 |
| 1986 | 0.095 | 0.138 | 0.164 | 0.194 | 0.212 | 0.225 | 0.239 | 0.208 | 0.288 |
| 1987 | 0.085 | 0.102 | 0.150 | 0.169 | 0.177 | 0.193 | 0.205 | 0.215 | 0.220 |
| 1988 |  | 0.098 | 0.133 | 0.153 | 0.166 | 0.171 | 0.183 | 0.191 | 0.201 |
| 1989 | 0.080 | 0.130 | 0.141 | 0.164 | 0.174 | 0.183 | 0.192 | 0.193 | 0.203 |
| 1990 | 0.094 | 0.138 | 0.148 | 0.160 | 0.176 | 0.189 | 0.194 | 0.208 | 0.216 |
| 1991 | 0.089 | 0.134 | 0.145 | 0.157 | 0.167 | 0.185 | 0.199 | 0.207 | 0.230 |
| 1992 | 0.095 | 0.141 | 0.147 | 0.157 | 0.165 | 0.171 | 0.180 | 0.194 | 0.219 |
| 1993 | 0.112 | 0.138 | 0.153 | 0.170 | 0.181 | 0.184 | 0.196 | 0.229 | 0.236 |
| 1994 | 0.081 | 0.141 | 0.164 | 0.177 | 0.189 | 0.187 | 0.191 | 0.204 | 0.220 |
| 1995 | 0.080 | 0.140 | 0.161 | 0.173 | 0.182 | 0.198 | 0.194 | 0.206 | 0.217 |
| 1996 | 0.085 | 0.135 | 0.172 | 0.182 | 0.199 | 0.209 | 0.220 | 0.233 | 0.237 |
| 1997 | 0.093 | 0.135 | 0.155 | 0.181 | 0.201 | 0.217 | 0.217 | 0.231 | 0.239 |
| 1998 | 0.095 | 0.136 | 0.145 | 0.173 | 0.191 | 0.196 | 0.202 | 0.222 | 0.217 |
| 1999 | 0.106 | 0.144 | 0.145 | 0.163 | 0.186 | 0.195 | 0.200 | 0.216 | 0.222 |
| 2000 | 0.102 | 0.129 | 0.154 | 0.172 | 0.180 | 0.184 | 0.204 | 0.203 | 0.204 |
| 2001 | 0.086 | 0.122 | 0.139 | 0.167 | 0.183 | 0.188 | 0.222 | 0.222 | 0.213 |
| 2002 | 0.097 | 0.127 | 0.140 | 0.155 | 0.175 | 0.196 | 0.204 | 0.218 | 0.226 |
| 2003 | 0.102 | 0.134 | 0.150 | 0.167 | 0.183 | 0.196 | 0.216 | 0.210 | 0.228 |
| 2004 | 0.085 | 0.140 | 0.150 | 0.167 | 0.182 | 0.193 | 0.222 | 0.221 | 0.285 |
| 2005 | 0.105 | 0.135 | 0.150 | 0.162 | 0.174 | 0.188 | 0.200 | 0.237 | 0.296 |
| 2006 | 0.106 | 0.137 | 0.141 | 0.158 | 0.169 | 0.178 | 0.199 | 0.221 | 0.243 |

Table 6.4.1.2. VIa(S) \& VIIb,c herring. Mean weight at age (winter rings) in the stock 1970 to 2006.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1971 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1972 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1973 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1974 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1975 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1976 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1977 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1978 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1979 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1980 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1981 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1982 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1983 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1984 | 0.120 | 0.169 | 0.210 | 0.236 | 0.260 | 0.273 | 0.283 | 0.290 | 0.296 |
| 1985 | 0.100 | 0.150 | 0.196 | 0.227 | 0.238 | 0.251 | 0.252 | 0.269 | 0.284 |
| 1986 | 0.098 | 0.169 | 0.209 | 0.238 | 0.256 | 0.276 | 0.280 | 0.287 | 0.312 |
| 1987 | 0.097 | 0.164 | 0.206 | 0.233 | 0.252 | 0.271 | 0.280 | 0.296 | 0.317 |
| 1988 | 0.097 | 0.164 | 0.206 | 0.233 | 0.252 | 0.271 | 0.280 | 0.296 | 0.317 |
| 1989 | 0.138 | 0.157 | 0.168 | 0.182 | 0.200 | 0.217 | 0.227 | 0.238 | 0.245 |
| 1990 | 0.113 | 0.152 | 0.170 | 0.180 | 0.200 | 0.217 | 0.225 | 0.233 | 0.255 |
| 1991 | 0.102 | 0.149 | 0.174 | 0.190 | 0.195 | 0.206 | 0.226 | 0.236 | 0.248 |
| 1992 | 0.102 | 0.144 | 0.167 | 0.182 | 0.194 | 0.197 | 0.214 | 0.218 | 0.242 |
| 1993 | 0.118 | 0.166 | 0.196 | 0.205 | 0.214 | 0.220 | 0.223 | 0.242 | 0.258 |
| 1994 | 0.098 | 0.156 | 0.192 | 0.209 | 0.216 | 0.223 | 0.226 | 0.230 | 0.247 |
| 1995 | 0.090 | 0.144 | 0.181 | 0.203 | 0.217 | 0.226 | 0.227 | 0.239 | 0.246 |
| 1996 | 0.086 | 0.137 | 0.186 | 0.206 | 0.219 | 0.234 | 0.233 | 0.249 | 0.253 |
| 1997 | 0.094 | 0.135 | 0.169 | 0.194 | 0.210 | 0.224 | 0.231 | 0.230 | 0.239 |
| 1998 | 0.095 | 0.136 | 0.145 | 0.173 | 0.191 | 0.196 | 0.202 | 0.222 | 0.217 |
| 1999 | 0.104 | 0.145 | 0.154 | 0.174 | 0.200 | 0.222 | 0.230 | 0.240 | 0.246 |
| 2000 | 0.100 | 0.134 | 0.157 | 0.177 | 0.197 | 0.207 | 0.217 | 0.230 | 0.245 |
| 2001 | 0.091 | 0.125 | 0.150 | 0.172 | 0.191 | 0.200 | 0.203 | 0.203 | 0.216 |
| 2002 | 0.092 | 0.127 | 0.146 | 0.170 | 0.190 | 0.201 | 0.210 | 0.227 | 0.229 |
| 2003 | 0.094 | 0.131 | 0.155 | 0.175 | 0.192 | 0.203 | 0.232 | 0.222 | 0.243 |
| 2004 | 0.081 | 0.133 | 0.151 | 0.175 | 0.194 | 0.207 | 0.238 | 0.233 | 0.276 |
| 2005 | 0.095 | 0.127 | 0.15 | 0.172 | 0.185 | 0.196 | 0.223 | 0.234 | 0.274 |
| 2006 | 0.092 | 0.130 | 0.133 | 0.162 | 0.177 | 0.186 | 0.209 | 0.238 | 0.247 |

Table 6.6.2.1. VIa(S) and VIIb,c herring VPA run with a terminal $F$ value of 0.2

|  | Traditional vpa Terminal populations from weighted Separable populations |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | RECRUITS Age 1 | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | SOPCOFAC | FBAR 3-6 |
| 1970 | 408071 | 207776 | 132321 | 20306 | 0.1535 | 0.8968 | 0.1754 |
| 1971 | 819057 | 226409 | 116628 | 15044 | 0.129 | 0.8707 | 0.1569 |
| 1972 | 736983 | 236280 | 122113 | 23474 | 0.1922 | 0.8975 | 0.1993 |
| 1973 | 537069 | 269391 | 156208 | 36719 | 0.2351 | 1.0162 | 0.2822 |
| 1974 | 594689 | 206306 | 93840 | 36589 | 0.3899 | 0.9762 | 0.4436 |
| 1975 | 410989 | 202278 | 100330 | 38764 | 0.3864 | 1.1237 | 0.4294 |
| 1976 | 691964 | 191879 | 69543 | 32767 | 0.4712 | 1.0472 | 0.4922 |
| 1977 | 585087 | 183576 | 78762 | 20567 | 0.2611 | 1.0778 | 0.3122 |
| 1978 | 1058130 | 228970 | 73952 | 19715 | 0.2666 | 1.0161 | 0.2573 |
| 1979 | 985833 | 268323 | 106500 | 22608 | 0.2123 | 1.0664 | 0.2652 |
| 1980 | 537870 | 204052 | 102250 | 30124 | 0.2946 | 0.9636 | 0.3842 |
| 1981 | 688029 | 221958 | 103766 | 24922 | 0.2402 | 1.0312 | 0.3056 |
| 1982 | 705717 | 232349 | 114538 | 19209 | 0.1677 | 1.0301 | 0.2201 |
| 1983 | 2329553 | 432025 | 109938 | 32988 | 0.3001 | 1.0042 | 0.3548 |
| 1984 | 971952 | 351102 | 185754 | 27450 | 0.1478 | 0.9688 | 0.2008 |
| 1985 | 1239791 | 358277 | 189803 | 23343 | 0.123 | 0.9846 | 0.1679 |
| 1986 | 952516 | 370399 | 224951 | 28785 | 0.128 | 0.9834 | 0.1777 |
| 1987 | 3259194 | 570764 | 196859 | 48600 | 0.2469 | 0.9488 | 0.3394 |
| 1988 | 483629 | 433212 | 304867 | 29100 | 0.0955 | 0.9992 | 0.2663 |
| 1989 | 718183 | 379758 | 227522 | 29210 | 0.1284 | 1.001 | 0.1789 |
| 1990 | 814900 | 345154 | 196803 | 43969 | 0.2234 | 1.0006 | 0.2558 |
| 1991 | 504653 | 272198 | 170015 | 37700 | 0.2217 | 0.9971 | 0.2403 |
| 1992 | 417747 | 219608 | 136119 | 31856 | 0.234 | 0.9951 | 0.2711 |
| 1993 | 617099 | 235622 | 117057 | 36763 | 0.3141 | 1.006 | 0.3509 |
| 1994 | 808342 | 216674 | 97624 | 33908 | 0.3473 | 0.998 | 0.3579 |
| 1995 | 469304 | 165722 | 85128 | 27792 | 0.3265 | 1.0525 | 0.4606 |
| 1996 | 839948 | 171639 | 63840 | 32534 | 0.5096 | 0.9955 | 0.5758 |
| 1997 | 833969 | 175635 | 65668 | 27225 | 0.4146 | 1.0016 | 0.5243 |
| 1998 | 540324 | 145104 | 53968 | 38895 | 0.7207 | 0.9988 | 0.9901 |
| 1999 | 408888 | 118142 | 46802 | 26109 | 0.5579 | 1.0018 | 0.6415 |
| 2000 | 497246 | 111116 | 39994 | 19846 | 0.4962 | 1.0011 | 0.4825 |
| 2001 | 538478 | 105801 | 39230 | 14756 | 0.3761 | 0.9988 | 0.574 |
| 2002 | 783099 | 133642 | 41089 | 17826 | 0.4338 | 0.9991 | 0.5939 |
| 2003 | 716622 | 144382 | 54941 | 16502 | 0.3004 | 1.002 | 0.4799 |
| 2004 | 966748 | 165637 | 66434 | 13727 | 0.2066 | 1.0006 | 0.3534 |
| 2005 | 677170 | 172553 | 83331 | 16231 | 0.1948 | 0.9986 | 0.2729 |
| 2006 | 719186* | 153260 | 83256 | 19193 | 0.2305 | 1.0012 | 0.2962 |
|  | metric Mean |  |  |  |  |  |  |

Table 6.6.2.2. VIa(S) and VIIbc herring VPA run using a terminal F or 0.4


Table 6.6.2.3 VIa(S) and VIIb,c herring VPA run using a terminal F or 0.6

| Traditional vpa Terminal populations from weighted Separable populations |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | RUITS | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | SOPCOFAC | FBAR 3-6 |
| Age 1 |  |  |  |  |  |  |  |
| 1970 | 411213 | 211932 | 135859 | 20306 | 0.1495 | 0.8968 | 0.1716 |
| 1971 | 825795 | 230537 | 119814 | 15044 | 0.1256 | 0.8707 | 0.1532 |
| 1972 | 744260 | 240598 | 125389 | 23474 | 0.1872 | 0.8975 | 0.1955 |
| 1973 | 543173 | 276043 | 161696 | 36719 | 0.2271 | 1.0162 | 0.2774 |
| 1974 | 601866 | 210080 | 96582 | 36589 | 0.3788 | 0.9762 | 0.4355 |
| 1975 | 417395 | 206593 | 103559 | 38764 | 0.3743 | 1.1237 | 0.4188 |
| 1976 | 703209 | 195933 | 72009 | 32767 | 0.455 | 1.0472 | 0.4777 |
| 1977 | 595724 | 188005 | 81539 | 20567 | 0.2522 | 1.0778 | 0.3014 |
| 1978 | 1080517 | 234688 | 76665 | 19715 | 0.2572 | 1.0161 | 0.2489 |
| 1979 | 1010739 | 275684 | 110234 | 22608 | 0.2051 | 1.0664 | 0.2557 |
| 1980 | 551172 | 210092 | 106288 | 30124 | 0.2834 | 0.9636 | 0.3693 |
| 1981 | 702747 | 229080 | 108607 | 24922 | 0.2295 | 1.0312 | 0.291 |
| 1982 | 721477 | 239869 | 119635 | 19209 | 0.1606 | 1.0301 | 0.2098 |
| 1983 | 2379992 | 444320 | 115660 | 32988 | 0.2852 | 1.0042 | 0.3389 |
| 1984 | 992740 | 362043 | 193359 | 27450 | 0.142 | 0.9688 | 0.1918 |
| 1985 | 1260704 | 368363 | 197037 | 23343 | 0.1185 | 0.9846 | 0.1607 |
| 1986 | 966100 | 380409 | 232889 | 28785 | 0.1236 | 0.9834 | 0.1709 |
| 1987 | 3298613 | 582680 | 204538 | 48600 | 0.2376 | 0.9488 | 0.3269 |
| 1988 | 488160 | 443343 | 313648 | 29100 | 0.0928 | 0.9992 | 0.2567 |
| 1989 | 722942 | 387565 | 234091 | 29210 | 0.1248 | 1.001 | 0.1733 |
| 1990 | 817816 | 351708 | 202611 | 43969 | 0.217 | 1.0006 | 0.2492 |
| 1991 | 505466 | 276937 | 174369 | 37700 | 0.2162 | 0.9971 | 0.2347 |
| 1992 | 417844 | 223298 | 139570 | 31856 | 0.2282 | 0.9951 | 0.267 |
| 1993 | 616928 | 238603 | 119901 | 36763 | 0.3066 | 1.006 | 0.3469 |
| 1994 | 806747 | 218847 | 99829 | 33908 | 0.3397 | 0.998 | 0.355 |
| 1995 | 465670 | 166095 | 85830 | 27792 | 0.3238 | 1.0525 | 0.458 |
| 1996 | 833589 | 171242 | 64019 | 32534 | 0.5082 | 0.9955 | 0.5752 |
| 1997 | 821778 | 174191 | 65437 | 27225 | 0.416 | 1.0016 | 0.5261 |
| 1998 | 526419 | 142771 | 53075 | 38895 | 0.7328 | 0.9988 | 1.0029 |
| 1999 | 387245 | 114179 | 45285 | 26109 | 0.5766 | 1.0018 | 0.664 |
| 2000 | 448025 | 103587 | 37678 | 19846 | 0.5267 | 1.0011 | 0.5092 |
| 2001 | 444829 | 92734 | 35226 | 14756 | 0.4189 | 0.9988 | 0.6294 |
| 2002 | 553642 | 104054 | 33634 | 17826 | 0.53 | 0.9991 | 0.6987 |
| 2003 | 418556 | 97271 | 38485 | 16502 | 0.4288 | 1.002 | 0.6367 |
| 2004 | 439292 | 91364 | 39031 | 13727 | 0.3517 | 1.0006 | 0.5645 |
| 2005 | 241132 | 78995 | 37806 | 16231 | 0.4293 | 0.9986 | 0.5454 |
| 2006 | 664351* | 55680 | 25297 | 19193 | 0.7587 | 1.0012 | 0.8829 |
|  | eometric M |  |  |  |  |  |  |



Figure 6.1.2.1. VIa(S) \& VIIb,c herring. Working group estimate of catches from 1970-2006.


Figure 6.1.3.1. VIa(S) \& VIIb,c herring. Northwest coast herring spawning grounds with arrows showing industry perceptions of herring movement.


Figure 6.3.2.1 VIa(S) \& Division VIIb,c herring. Cruise track and trawl positions during the 2007 northwest herring acoustic survey.


Figure 6.3.2.2 VIa(S) \& Division VIIb,c herring. Post plot showing the distribution of total herring SA values obtained during the 2007 northwest herring acoustic survey.



Figure 6.3.2.3. VIa(S) \& Division VIIb,c herring. Age (winter rings) distributions of the abundance estimate from the 2006 and 2007 acoustic surveys and of the fishery in 2006 (Above). Age distribution of the abundance estimates from 4 acoustic surveys; 2004, 2005, 2006 and 2007 (below).


Figure 6.4.4.1. VIa(S) \& Division VIIb,c herring. Mean weight in the catch 1982 - 2006.


Figure 6.4.4.2. VIa(S) \& Division VIIb,c herring. Mean weight in the stock 1982 - 2006.



Figure 6.6.1.1 VIa(S) \& Division VIIb,c herring. Mean standardised catch numbers at age standardised by year for the fishery (above) and the survey (below)



Figure 6.6.1.2. VIa (S) and VIIb,c herring. Log catch ratios by year (upper) and log catch ratios with a four-year running average (lower).


Figure 6.6.1.3: VIa(S) and VIIb,c herring. Cohort catch curves for the time series of catch at age data.


Figure 6.6.1.4: VIa(S) and VIIb,c herring. Mean log catch numbers at age for ages $\mathbf{3 - 7}$ from the fishery and mean abundance estimates at age from the acoustic survey.


Figure 6.6.2.1. VIa(S) and VIIb,c herring comparison of three separable VPA runs of the current working group and the 2006 working group, using values of $0.2,0.4$ and 0.6 for terminal $F$.




Figure 6.6.2.2. VIa(S) and VIIb,c herring - Residuals from three separable VPA runs using terminal $F$ values of 0.2 (upper), 0.4 (middle) and 0.6 (lower). Black indicates positive residuals and white indicates negative.




Figure 6.6.2.3. VIa(S) and VIIb,c herring - Results from two ICA runs and the three separable VPA runs.


Figure 6.6.2.4. VIa(S) and VIIb,c herring - Separable model and survey residual patterns from ICA runs including and excluding the 2007 acoustic survey.


Figure 6.6.2.5. VIa(S) and VIIb,c herring - Year and age residual patterns from ICA runs including and excluding the 2007 acoustic survey.


Figure 6.6.2.6. VIa(S) and VIIb,c herring. Scatter plot of estimates of F and SSB for the terminal year using parameter estimate variance-covariance matrix estimates in a bootstrap evaluation of the precision of the assessments, both including and excluding the 2007 acoustic survey.

## $7 \quad$ Irish Sea Herring [Division VIIA(North)]

### 7.1 The Fishery

### 7.1.1 Advice and Management Applicable to 2006 and 2007

The WG did not present the results of a final assessment to ACFM in 2006 due to uncertainties in the estimates of F and SSB observed in the exploratory ICA runs. The exploratory analysis did suggest however that the SSB has been relatively stable for the last 10 years, and fishing mortality has not increased above the recent average.

ACFM subsequently advised that a status quo TAC of 4800 t be adopted for 2006, partitioned as 3500 t to the UK and 1250 t to the Republic of Ireland.

Closed areas for herring fishing in the Irish Sea along the east coast of Ireland and within 12 nautical miles of the west coast of Britain were maintained throughout the year. The traditional gillnet fishery on the Mourne herring, which has a derogation to fish within the Irish closed box, operated successfully in 2006, having returned in 2005 after many years of absence. The area to the east of the Isle of Man, encompassing the Douglas Bank spawning ground (described in ICES 2001, ACFM:10), was closed from $21^{\text {st }}$ September to $15^{\text {th }}$ November. Boats from the Republic of Ireland are not permitted to fish east of the Isle of Man.

### 7.1.2 The Fishery in 2006

The catches reported from each country for the period 1986 to 2006 are given in Table 7.1.1, and total catches from 1961 to 2006 in Figure 7.1.1. Reported international landings in 2006 for the Irish Sea amounted to 4402 t with UK vessels acquiring extra quota through swaps with the Republic of Ireland. As with recent years the majority of the catch of herring in VIIa( N ) was taken during the $3^{\text {rd }}$ quarter.

The 2006 VIIa(N) herring fishery opened in August, with activity based mainly to the west of the Isle of Man. The majority of catches taken during August and September were by a pair of UK pair trawlers. One pair of polyvalent RSW Republic of Ireland vessels also took part in the fishery. Catches were reported from the last week in August to the third week in September.

September saw the opening of the Mourne fishery, limited to boats under 40ft, fishing with driftnets. This fishery saw reasonable catches of herring in 2005 for the first time in many years and again operated in 2006 landing catches of $\sim 20 \mathrm{t}$ during October and November. Fishing by the UK pair trawlers recommenced in November and continued through to December.

### 7.2 Biological composition of the Catch

### 7.2.1 Catch in numbers

Catches in numbers-at-age are given in Table 7.2 .1 for the years 1972 to 2006 and a graphical representation is given in Figure 7.2.1. The predominant year class in 2006 landings was the 2-ringers (2003-year class) with the highest catch seen since 1999. The large numbers of 3ringers observed in 2005 catches continue to be detected in 2006 catches as 4 -ringers. The catch in numbers at length is given in Table 7.2.2 for 1991 to 2006.

### 7.2.2 Quality of catch and biological data

There are no estimates of discarding or slippage in the Irish Sea fisheries that target herring. Discarding however is not thought to be a feature of this fishery. Biological sampling remains high for this fishery with data arising from both the Republic of Ireland and Northern Ireland laboratories. It should be noted however that the majority of samples are taken from only one fishing unit, the pair of UK vessels operating in the Irish Sea. 20 samples were processed from the $3^{\text {rd }}$ quarter fishery and 2 from the driftnet fishery operating during the $4^{\text {th }}$ quarter (Table 7.2.3).

### 7.2.3 Acoustic surveys

The information on the time-series of acoustic surveys in the Irish Sea is given in Table 7.2.4. As in last year's assessment, the SSB estimates from the survey are calculated using the same (annually varying) maturity ogives that are input to ICA (see Table 7.3.3 estimated from the commercial catch data).

The acoustic survey in 2006 was carried out over 11 days in the period $30^{\text {th }}$ August to $9^{\text {th }}$ September. A survey design of stratified, systematic transects was employed, as in previous years (Figure 7.2.2.A). In general, there are few samples on the age composition of the herring in the acoustic survey data. The survey followed the methods described in Armstrong et al., 2005 WD 23; (see Annex 2). Sampling intensity was high during the 2006 survey with 28 successful trawls completed. The length frequencies generated from these trawls highlights the spatial heterogeneous nature of herring age groups in the Irish Sea (Figure 7.2.4)

The bulk of the acoustic scatter attributed to pelagic fish was identified as sprat, which were abundant around the periphery of the Irish Sea and to the west of the Isle of Man (Figure 7.2.2.B). 0 -group herring were found to be abundant to the west of the Isle of Man and in the north eastern Irish Sea (Figure 7.2.3.B). Mixed 1+ herring targets were scattered evenly around the Isle of Man, with larger shoals off the north western coast (Figure 7.2.3.A). Comparing the mixed herring distributions with 2005 highlighted the shift in biomass distribution from the east to the west Isle of Man. This shift in biomass between the eastern and western coasts is a pattern observed throughout the acoustic time-series and is thought to be associated with the timing of migration to the spawning grounds.

As in previous years, no herring schools were detected in the area immediately north of the Isle of Man, despite an abundance of early-stage larvae in this area in November (Figure 7.2.5). It is possible that spawning in this area only commences after the date of the acoustic survey.

The estimate of herring SSB of 16332 t for 2006 is a marked reduction on last years estimate (Table 7.2.4). The approximate coefficient of variation (CV) of 0.22 is low due in part to the scattered nature of the mixed herring targets detected during the survey. The biomass estimate of 33136 t for $1+$ ringers is also a reduction on last year, whilst the approximate CV of 0.24 is an improvement. Given the approximate CVs of the estimates, it is not possible to discern any trend in 1+ biomass or SSB since 1999.

The estimate of the herring population, excluding 0 -ring fish, is given in Table 7.2.5. The age composition from the acoustic survey was similar to the catch-at-age data with higher proportion catches of 2 and 3-ringers.

### 7.2.4 Larvae surveys

Northern Ireland undertook a herring larvae survey over the period $6^{\text {th }}$ to $10^{\text {th }}$ November 2006. The survey followed the methods and designs of previous surveys in the time-series (see Annex 2). The production estimate for 2006 in the NE Irish Sea was the third highest in the
time series and well above the series average (Table 7.2.6). Herring larvae were found to be most abundant to the southeast and northeast of the Isle of Man and less abundant in the western Irish Sea (Figure 7.2.4).

### 7.2.5 Groundfish surveys of Area VIIa(N)

Groundfish surveys (Annex 2), carried out by Northern Ireland since 1991 in the Irish Sea, were used by the 1996 to 1999 HAWG to obtain indices for 0- and 1-ring herring in the Irish Sea. These indices have performed poorly in the assessment and have not been used since 1999. New values were added to the time-series in 2006 and are shown in Table 7.2.7.

### 7.3 Mean length, weight, maturity and natural mortality-at-age

Mean lengths-at-age were calculated using the catch data and are given for the years 1985 to 2006 in Table 7.3.1. In general, mean lengths have been relatively stable over the last few years and this trend has continued in 2006.

Mean weights-at-age in the catch are given in Table 7.3.2. Mean weights-at-age of all ages remained low. There has been a change in mean weight over the time period 1961 to the present (ICES 2003 ACFM:17). Mean weights-at-age increased between the early 1960s and the late 1970s whereupon there has been a steady decline to the early 1990s, where they remained low. In the assessment, mean weights-at-age for the period 1972 to 1984 are taken as unchanging. In extending the data series back from 1971 to 1961, mean weights-at-age in the catch were taken from samples recorded by the Port Erin Marine Laboratory (ICES 2003 ACFM:17).

Mean weights-at-age in the third-quarter catches (for the whole time-series 1961 to present) have been used as estimates of stock weights at spawning time. There was some uncertainty in the mean weights-at-age for 2003 presented to the WG, and consequently the WG replaced these with the average mean stock weights-at-age for the preceding five years (1998 to 2002).

Maturity-at-age (in the catches) for each year (1961 to 2005) are given in Table 7.3.3. Due to inconsistencies in the maturity data collected in 2003, the WG used a mean maturity ogive for the preceding nine years for 2003. The rationale for the 9 years was that there appeared to be a shift in the maturity ogive around 1993. After 2003 all weights and maturity-at-age data were based on corresponding annual biological samples.

As in previous years, natural mortality per year was assumed to be 1.0 on 1-ringers, 0.3 for 2ringers, 0.2 for 3 -ringers and 0.1 for all older age classes. These are based on the natural mortality rates estimated for herring in the North Sea using MSVPA.

### 7.4 Recruitment

An estimate of total abundance of 1-ringers is provided by the Northern Ireland acoustic survey. However, there is evidence that a fraction of those is of Celtic Sea origin. Separation of the trawl catches of juveniles into autumn and winter spawning components, based on otolith microstructure and/or length composition, could result in a survey index of recruitment for the Irish Sea stock that could be used directly in the assessment. Such an index may also be of use in the Celtic Sea assessment, as it would provide an estimate of juveniles resident in the Irish Sea originating from this management area.

### 7.5 Stock Assessment

### 7.5.1 Data exploration and preliminary modelling

In 2006 two fishery independent survey indices were used as tuning indices to run ICA: Northern Irish larvae production (NINEL) and the age dis-aggregated abundance index from the acoustic survey (ACAGE). The preliminary modelling used catch-at-age data derived from the landings, extending back to 1961.

2006 data were added to the Northern Irish larvae series (NINEL), the Northern Irish acoustic survey (total biomass, SSB and age-structured indices) and the catch-at-age data derived from the landings. Due to the continuing problems associated with mixing of Irish Sea and Celtic Sea juveniles the groundfish surveys were considered unsuitable tuning fleets and are not considered further. The survey series available for inclusion in an assessment using the ICA package are documented in Appendix 2.

Initial fits within integrated catch-at-age analysis (ICA), were performed with NINEL and ACAGE. The following model settings were used:

Separable constraint over the last 6-years (weighting = 1.0 for each year)
Reference age $=4$
Constant selection pattern model
Selectivity on oldest age $=1.0$
First age for calculation of mean $\mathrm{F}=2$
Last age for calculation of mean $\mathrm{F}=6$
Weighting on 1-rings $=0.1$; all other age classes $=1.0$
Weighting for all years $=1.0$
All indices treated as linear
No S/R relationship fitted
Lowest and highest feasible F $=0.05$ and 2.0
All survey weights fitted by hand i.e., 1.0 with the 1 -rings in the acoustic survey weighted to 0.1 .
Correlated errors assumed i.e., $=1.0$
No shrinkage applied
The initial fit corresponds to the same procedure as last year (SPALY).
Examination of the initial fit SSQ surface (Figure 7.5.1) showed large discrepancies in the overall minima of reference $F$ for the NINEL tuning index and ACAGE index. The NINEL tuning index showed a minimum at between 0.36 and 0.56 while the ACAGE index returned no readily discernible minimum. This highlighted the contrasting effect of these separate tuning indices on the deterministic calculation of the 6-year separable period. Comparison of reference Fs estimated in the SPALY run and ICA runs using one index at a time (NINEL, ACAGE) highlighted the variation in mean F (2-6) and associated deviation (Figure 7.5.2). The NINEL index indicated the lowest reference $F(0.246)$ compared to the acoustic index (1.847), with the SPALY run indicating a reference $F$ (1.177) intermediate between the two. The ACAGE reference $F$ was estimated with very wide confidence intervals reflected also in the SPALY estimate.

Historical trends in log catch ratios along cohorts were investigated to give some indication of trends in Z over the time-series (Figure 7.5.3). The mean depletion rates of cohorts show a very high rate during the 1975 to 79 year classes (Figure 7.5.4). This period corresponds to a time of intensive fishing activity and associated high landings in the Irish Sea. Since the 1980 to 84 year classes the data suggest that total mortality has been slowly increasing. Estimates
since 2000 shown a further decline in Z but are not accurate as the recruiting year classes since then have not passed fully through the fishery.

Inter-annual variation in the proportion catch at age (\%) estimated from catch at age and the acoustic survey was explored to consider the signal generated by these data (Figure 7.5.5). Strong year and age effects were shown to occur throughout both time-series at all age classes (rings). Increases in proportion catches at age 2 were shown over the time-series in the acoustic surveys, while at age 8 a downward trend is seen from 1994 to present in both the catch and acoustic data. Age 1 were not included in the analysis due to the possible mixing effects with Celtic Sea fish and the reduced selectivity to the fishery.

Year effects were evident in every age class and seemed to be consistent between the acoustic and catch data during the early part of the time-series, however during 2000/2001 this consistency became uncoupled. This seems not the case for the age 2 class however, were an apparent contrasting pattern between the catch and acoustic survey is evident throughout the time- series. Comparing the proportion catch-at-age data with the ICA age and year residuals in the separable period it can be seen that the year effect from 2003 is evident in both. (Figure 7.5.9).

The year effects present in the age 2 catch and acoustic data were investigated further through examination of the spatial distribution of adult herring (1+ring) biomass during the acoustic survey. The shift in biomass between the eastern and western coasts surrounding the Isle of Man is an anomaly observed throughout the acoustic time-series and is thought to be associated with the timing of migration to the spawning grounds (Figure 7.5.6). An oscillating pattern of biomass distribution is observed where the bulk of adult herring biomass is either found on the western or eastern coast of the Isle of Man. A significant correlation of the annual ratio of the proportion in catch at age 2 between the acoustic and catch data, and the annual western adult herring (1+) biomass estimate was found (Figure 7.5.7). This correlation suggests that the timing of migration and distribution of herring have a significant effect on the resulting catch at age 2 arising from the fishery. As the landings from this fishery mainly arise from 1 operational unit it is not surprising that strong year effects are seen in the catch-at-age data. This will have major implications for the precision of the assessment in $\mathrm{VIIa}(\mathrm{N})$ and may explain the contrasting signal between the catch at age and acoustic tuning index.

## Two-stage biomass model

In 2005 a Two-Stage Biomass model for the assessment of Irish Sea VIIa herring given additional variance in the recruitment index was presented by Roel and De Oliveira (2005 WD10). In 2006 due to the uncertainty in the outputs from the SPALY ICA run and the mixing of 0 -group Irish and Celtic Sea recruits in the management area, the model was not attempted at the 2006 HAWG.

### 7.5.2 Conclusion to explorations

The results from the exploratory runs carried out with ICA using NINEL and ACAGE as tuning indices indicate a low precision in reference $\mathrm{F}(2-6)$ in 2006 (Figure 7.5.2). Exploration of proportion at age data suggests that conflicting year effects are present in the acoustic and catch at age data, particularly since 2000/2001 (Figure 7.5.5). These conflicting signals are contributing to the poor model fit in the separable period as shown by the year residuals generated from ICA (Figure 7.5.9). These year effects therefore may also explain the low precision in reference $F$ generated by the model.

It is evident from the exploratory analysis that there is a contrast in the proportion catch at age between the acoustic and catch at age data. There is evidence that the inter-annual variation in the migration and distribution of herring surrounding the Isle of Man has an inter-annual effect on the age class selectivity of this fishery. This effect is shown to be significantly associated
with the age 2 year classes, the most abundant in numbers at age in the stock (Table 7.2.1). This violates the separable assumption of ICA.

Further investigation into a more robust assessment model for this stock that does not assume separability is required.

### 7.5.3 Stock Assessment

The results presented correspond to SPALY ICA runs using the acoustics data as an agestructured index (ACAGE) and the Northern Ireland larval survey (NINEL) as an index of biomass (Figure 7.5.8-7.5.10). The outputs shown are an updated exploratory assessment and no analytical assessment is presented this year. The model settings are the same as for 2005. The run log for the assessment is shown in Table 7.5.1. The output from ICA assessment, the residuals and fitted values are given in Tables 7.5.2-7.5.19. This assessment provides information on the relative trends of the converged period of the VPA for this stock. However it does not provide reliable information on the period covered by the 6 -year separable period.

### 7.6 Stock and Catch Projection

### 7.6.1 Deterministic short-term predictions

No short-term predictions were included in the 2006 assessment.

### 7.6.2 Yield-per-recruit

For a yield-per-recruit analysis refer to last years report.

### 7.7 Medium-term predictions of stock size

The Working Group decided that there was no basis for undertaking medium-term projections of stock size until there is agreement that advice based on the assessment can be provided.

### 7.8 Reference points

The estimation of $\mathbf{B}_{\mathrm{pa}}(9500 \mathrm{t})$ and $\mathbf{B}_{\mathrm{lim}}$ (6000t) were not revisited this year. There were no new points to add to the discussions and deliberations presented in 2000 (ICES 2000/ACFM:12). There is no precautionary F value for this stock.

### 7.9 Quality of the Assessment

The different survey series for Irish Sea herring are characterised by generally poor precision caused by the very patchy distribution of the fish as well as assumptions inherent in the methods (e.g. target strength, larval growth and mortality; relationship between larval production and SSB, constant selectivity in the separable period). Nonetheless, there is evidence of some coherence between the longer-term signals in the different survey series. The acoustic survey provides estimates of abundance at age but the juveniles in the area are a mixture of at least two adjacent stocks (Celtic Sea and VIIa(N)). Separation of trawl catches of juveniles into autumn and winter spawning components, based on otolith microstructure and/or length composition, could result in acoustic and trawl survey indices of juveniles appropriate for the Irish Sea assessment. However information from historical herring larval surveys in the Celtic Sea suggest that a considerable autumn spawning component existed in certain years. This may undermine attempts to separate the Celtic and Irish Sea juvenile components.

Retrospective analysis of the assessment with data from 2004 to 2006 was carried out (Figure 7.5.11). The retrospectives for SSB and $\mathrm{F}_{2-6}$ from the ICA assessment (NINEL + ACAGE
show instability in the estimation of SSB during the 6-year separable period. The tendency to over-estimate SSB in the last assessment year was shown to be a continuing trend. The retrospective pattern of $\mathrm{F}_{2-6}$ is relatively stable, however the addition of data in 2006 had the effect to revise F upwards. As this was an exploratory assessment these data should be taken as an exploratory exercise. The estimation of recruitment in the final year is not considered reliable (Figure 7.5.11). There is a systematic bias in the retrospective pattern arising from ICA.

For many years, the assessment for this stock has not been accepted by ACFM. Both the catches and survey data are noisy. From the exploratory analysis it can be seen that some of this noise may arise from the apparent inter-annual variation in herring migration patterns.

Given the noise in the data it is difficult to detect abrupt changes in the stock dynamics. Nevertheless some inferences can be made that are quite robust, even though the absolute estimates of SSB and fishing mortality may be less reliable. In particular, it seems likely that the stock is relatively stable at a level close to Bpa, and that the fishing mortality has been relatively stable since the late 1990's.

### 7.10 Spawning and Juvenile Fishing Area Closures

The arrangement of closed areas in Division VIIa(N) prior to 1999 are discussed in detail in ICES (1996/ACFM:10) with a change to the closed area to the east of the Isle of Man being altered in 1999 (ICES 2001/ACFM:10). The closed areas consist of: all year juvenile closures along part of the east coast of Ireland, and the west coast of Scotland, England and Wales; spawning closures along the east coast of the Isle of Man from 21st September- 15th November, and along the east coast of Ireland all year round. The WG recommends that any alterations to the present closures be considered carefully, in the context of this report, to ensure protection for all components of this stock.

### 7.11 Management considerations

The catches have been low in recent years and the fishing activity has not varied considerably as shown from landing data (Figure 7.1.1). There is evidence of a contraction in the age structure of this stock in both the catch and survey data (Figure 7.5.5). A further reduction in precision was noted in the 2006 assessment with the SSB estimated to be below $\mathrm{B}_{\mathrm{pa}}$ and above $\mathrm{B}_{\text {lim. }}$. Analytical retrospectives show considerable downward revision of SSB in subsequent assessments in recent years, placing SSB below the $\mathrm{B}_{\mathrm{pa}}$. Though the exact level of the stock is unclear from the current assessment model the trends indicate that SSB remains relatively stable. Recruitment is approximately average for the period since the 1980s when a change in the productivity of this stock was observed (Section 1.8.3). Therefore, the maintenance of recommended catch levels at current levels 4800 t , in the short-term, should not be detrimental to the stock.

A review of the model (ICA) currently employed in the assessment of this stock is recommended in light of the inter-annual variation in age class selectivity of the fishery.

Table 7.1.1 Irish Sea Herring Division VIIa(N). Working group catch estimates in tonnes by country, 1987-2006. The total catch does not in all cases correspond to the official statistics and cannot be used for management purposes.

| COUNTRY | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ireland | 1200 | 2579 | 1430 | 1699 | 80 | 406 | 0 | 0 | 0 |
| UK | 3290 | 7593 | 3532 | 4613 | 4318 | 4864 | 4408 | 4828 | 5076 |
| Unallocated | 1333 | - | - | - | - | - | - | - | - |
| Total | 5823 | 10172 | 4962 | 6312 | 4398 | 5270 | 4408 | 4828 | 5076 |
| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| Ireland | 100 | 0 | 0 | 0 | 0 | 862 | 286 | 0 | 749 |
| UK | 5180 | 6651 | 4905 | 4127 | 2002 | 4599 | 2107 | 2399 | 1782 |
| Unallocated | 22 | - | - | - | - | - |  | - | - |
| Total | 5302 | 6651 | 4905 | 4127 | 2002 | 5461 | 2393 | 2399 | 2531 |
| Country | 2005 | 2006 |  |  |  |  |  |  |  |
| Ireland | 1153 | 581 |  |  |  |  |  |  |  |
| UK | 3234 | 3821 |  |  |  |  |  |  |  |
| Unallocated | - | - |  |  |  |  |  |  |  |
| Total | 4387 | 4402 |  |  |  |  |  |  |  |

Table 7.2.1 Irish Sea Herring Division VIIa(N). Catch in numbers (thousands) by year.

|  | Age (rings) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| 1972 | 40640 | 46660 | 26950 | 13180 | 13750 | 6760 | 2660 | 1670 |
| 1973 | 42150 | 32740 | 38240 | 11490 | 6920 | 5070 | 2590 | 2600 |
| 1974 | 43250 | 109550 | 39750 | 24510 | 10650 | 4990 | 5150 | 1630 |
| 1975 | 33330 | 48240 | 39410 | 10840 | 7870 | 4210 | 2090 | 1640 |
| 1976 | 34740 | 56160 | 20780 | 15220 | 4580 | 2810 | 2420 | 1270 |
| 1977 | 30280 | 39040 | 22690 | 6750 | 4520 | 1460 | 910 | 1120 |
| 1978 | 15540 | 36950 | 13410 | 6780 | 1740 | 1340 | 670 | 350 |
| 1979 | 11770 | 38270 | 23490 | 4250 | 2200 | 1050 | 400 | 290 |
| 1980 | 5840 | 25760 | 19510 | 8520 | 1980 | 910 | 360 | 230 |
| 1981 | 5050 | 15790 | 3200 | 2790 | 2300 | 330 | 290 | 240 |
| 1982 | 5100 | 16030 | 5670 | 2150 | 330 | 1110 | 140 | 380 |
| 1983 | 1305 | 12162 | 5598 | 2820 | 445 | 484 | 255 | 59 |
| 1984 | 1168 | 8424 | 7237 | 3841 | 2221 | 380 | 229 | 479 |
| 1985 | 2429 | 10050 | 17336 | 13287 | 7206 | 2651 | 667 | 724 |
| 1986 | 4491 | 15266 | 7462 | 8550 | 4528 | 3198 | 1464 | 877 |
| 1987 | 2225 | 12981 | 6146 | 2998 | 4180 | 2777 | 2328 | 1671 |
| 1988 | 2607 | 21250 | 13343 | 7159 | 4610 | 5084 | 3232 | 4213 |
| 1989 | 1156 | 6385 | 12039 | 4708 | 1876 | 1255 | 1559 | 1956 |
| 1990 | 2313 | 12835 | 5726 | 9697 | 3598 | 1661 | 1042 | 1615 |
| 1991 | 1999 | 9754 | 6743 | 2833 | 5068 | 1493 | 719 | 815 |
| 1992 | 12145 | 6885 | 6744 | 6690 | 3256 | 5122 | 1036 | 392 |
| 1993 | 646 | 14636 | 3008 | 3017 | 2903 | 1606 | 2181 | 848 |
| 1994 | 1970 | 7002 | 12165 | 1826 | 2566 | 2104 | 1278 | 1991 |
| 1995 | 3204 | 21330 | 3391 | 5269 | 1199 | 1154 | 926 | 1452 |
| 1996 | 5335 | 17529 | 9761 | 1160 | 3603 | 780 | 961 | 1364 |
| 1997 | 9551 | 21387 | 7562 | 7341 | 1641 | 2281 | 840 | 1432 |
| 1998 | 3069 | 11879 | 3875 | 4450 | 6674 | 1030 | 2049 | 451 |
| 1999 | 1810 | 16929 | 5936 | 1566 | 1477 | 1989 | 444 | 622 |
| 2000 | 1221 | 3743 | 5873 | 2065 | 558 | 347 | 251 | 147 |
| 2001 | 2713 | 11473 | 7151 | 13050 | 3386 | 936 | 650 | 803 |
| 2002 | 179 | 9021 | 1894 | 1866 | 2395 | 953 | 474 | 343 |
| 2003 | 694 | 4694 | 3345 | 2559 | 882 | 2945 | 872 | 605 |
| 2004 | 3225 | 8833 | 5405 | 2161 | 623 | 213 | 673 | 127 |
| 2005 | 8692 | 13980 | 10555 | 3287 | 1422 | 415 | 292 | 368 |
| 2006 | 5669 | 15253 | 198 | 6318 | 1325 | 605 | 262 | 246 |

Table 7.2.2 Irish Sea Herring Division VIIa(N). Catch-at-length data 1991-2006. Numbers of fish in thousands. Table amended with 1990 year-class removed, see 2005 report.

| Length | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 14.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 15 |  | 95 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 15.5 |  | 169 |  |  |  |  |  |  | 10 |  |  |  |  |  |  |  |
| 16 |  | 343 |  |  | 21 | 21 | 17 |  | 19 | 12 | 9 |  |  |  |  | 2 |
| 16.5 | 2 | 275 |  |  | 55 | 51 | 94 |  | 53 | 49 | 27 |  |  | 13 | 1 | 44 |
| 17 | 1 | 779 |  | 84 | 139 | 127 | 281 | 26 | 97 | 67 | 53 |  |  | 25 | 39 | 140 |
| 17.5 | 4 | 1106 |  | 59 | 148 | 200 | 525 | 30 | 82 | 97 | 105 |  |  | 84 | 117 | 211 |
| 18 | 31 | 1263 |  | 69 | 300 | 173 | 1022 | 123 | 145 | 115 | 229 |  |  | 102 | 291 | 586 |
| 18.5 | 56 | 1662 |  | 89 | 280 | 415 | 1066 | 206 | 135 | 134 | 240 | 36 |  | 114 | 521 | 726 |
| 19 | 168 | 1767 | 39 | 226 | 310 | 554 | 1720 | 317 | 234 | 164 | 385 | 18 |  | 203 | 758 | 895 |
| 19.5 | 174 | 1189 | 75 | 241 | 305 | 652 | 1263 | 277 | 82 | 97 | 439 | 0 | 29 | 269 | 933 | 1246 |
| 20 | 454 | 1268 | 75 | 253 | 326 | 749 | 1366 | 427 | 218 | 109 | 523 | 0 | 73 | 368 | 943 | 984 |
| 20.5 | 341 | 705 | 57 | 270 | 404 | 867 | 1029 | 297 | 242 | 85 | 608 | 18 | 215 | 444 | 923 | 1443 |
| 21 | 469 | 705 | 130 | 400 | 468 | 886 | 1510 | 522 | 449 | 115 | 1086 | 307 | 272 | 862 | 1256 | 1521 |
| 21.5 | 296 | 597 | 263 | 308 | 782 | 1258 | 1192 | 549 | 362 | 138 | 1201 | 433 | 290 | 1007 | 1380 | 1621 |
| 22 | 438 | 664 | 610 | 700 | 1509 | 1530 | 2607 | 1354 | 1261 | 289 | 1748 | 1750 | 463 | 1495 | 1361 | 2748 |
| 22.5 | 782 | 927 | 1224 | 785 | 2541 | 2190 | 2482 | 1099 | 2305 | 418 | 1763 | 1949 | 600 | 2140 | 1448 | 3629 |
| 23 | 1790 | 1653 | 2016 | 1035 | 4198 | 2362 | 3508 | 2493 | 4784 | 607 | 2670 | 2490 | 1158 | 2089 | 1035 | 4358 |
| 23.5 | 1974 | 1156 | 2368 | 1473 | 4547 | 2917 | 3902 | 2041 | 4183 | 951 | 2254 | 1552 | 1380 | 2214 | 1256 | 2920 |
| 24 | 2842 | 1575 | 2895 | 2126 | 4416 | 3649 | 4714 | 3695 | 4165 | 1436 | 3489 | 1029 | 1273 | 2054 | 1276 | 3679 |
| 24.5 | 2311 | 2412 | 2616 | 2564 | 3391 | 4077 | 4138 | 2769 | 3397 | 1783 | 4098 | 758 | 1249 | 2269 | 1083 | 2431 |
| 25 | 2734 | 2792 | 2207 | 3315 | 3100 | 4015 | 5031 | 2625 | 2620 | 2144 | 5566 | 776 | 1163 | 1749 | 1086 | 3438 |
| 25.5 | 2596 | 3268 | 2198 | 3382 | 2358 | 3668 | 3971 | 2797 | 1817 | 1791 | 4785 | 1335 | 1211 | 1206 | 584 | 2198 |
| 26 | 3278 | 3865 | 2216 | 3480 | 2334 | 2480 | 3871 | 3115 | 1694 | 1349 | 3814 | 1570 | 1140 | 823 | 438 | 1714 |
| 26.5 | 2862 | 3908 | 2176 | 2617 | 1807 | 2177 | 2455 | 2641 | 1547 | 840 | 2243 | 1552 | 1573 | 587 | 203 | 605 |
| 27 | 2412 | 3389 | 2299 | 2391 | 1622 | 1949 | 1711 | 2992 | 1475 | 616 | 1489 | 776 | 1607 | 510 | 165 | 445 |
| 27.5 | 1449 | 2203 | 2047 | 1777 | 990 | 1267 | 1131 | 1747 | 867 | 479 | 644 | 433 | 1189 | 383 | 60 | 155 |
| 28 | 922 | 1440 | 1538 | 1294 | 834 | 906 | 638 | 1235 | 276 | 212 | 496 | 162 | 726 | 198 | 45 | 104 |
| 28.5 | 423 | 569 | 944 | 900 | 123 | 564 | 440 | 170 | 169 | 58 | 179 | 108 | 569 | 51 | 18 | 9 |
| 29 | 293 | 278 | 473 | 417 | 248 | 210 | 280 | 111 | 61 | 42 | 10 | 36 | 163 |  | 12 | 46 |
| 29.5 | 129 | 96 | 160 | 165 | 56 | 79 | 59 | 92 |  | 12 | 0 | 36 | 129 |  |  |  |
| 30 | 82 | 70 | 83 | 9 | 40 | 32 | 8 | 84 |  | 6 | 9 |  | 43 |  |  |  |
| 30.5 | 36 | 36 | 15 | 27 | 5 | 0 | 5 | 3 |  |  |  |  | 43 |  |  |  |
| 31 | 12 | 2 | 4 |  | 1 | 2 |  |  |  |  |  |  | 43 |  |  |  |
| 31.5 | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 32 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 32.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 33 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 33.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 34 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 7.2.3 Irish Sea Herring Division VIIa(N). Sampling intensity of commercial landings in 2006.

| Quarter | Country | LANDINGS <br> (T) | No. SAMPLES | No. FISH MEASURED | $\begin{gathered} \text { No. FISH } \\ \text { AGED } \\ \hline \end{gathered}$ | Estimation of discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Ireland | 0 | - | - | - | - |
|  | UK (N. Ireland) | 0.03 | 0 | 0 | 0 | No |
|  | UK (Isle of Man) | 0 | - | - | - | - |
|  | UK (Scotland) | 0 | - | - | - | - |
|  | UK (England \& Wales) | 0 | - | - | - | - |
| 2 | Ireland | 0 | - | - | - | - |
|  | UK (N. Ireland) | 0 | - | - | - | - |
|  | UK (Isle of Man) | * | - | - | - | - |
|  | UK (Scotland) | 0 | - | - | - | - |
|  | UK (England \& Wales) | 0 | - | - | - | - |
| 3 | Ireland | 581 | 8 | 2248 | 549 | No |
|  | UK (N. Ireland) | 3307 | 12 | 1406 | 586 | No |
|  | UK (Isle of Man) | * | - | - | - | - |
|  | UK (Scotland) | 0 | - | - | - | - |
|  | UK (England \& Wales) | 0 | - | - | - | - |
| 4 | Ireland | 0 | - | - | - | - |
|  | UK (N. Ireland) | 514 | 2 | 576 | 100 | No |
|  | UK (Isle of Man) | * | - | - | - | - |
|  | UK (Scotland) | 0 | - | - | - | - |
|  | UK (England \& Wales) | 0 | - | - | - | - |

[^10]Table 7.2.4 Irish Sea Herring Division VIIa(N). Summary of acoustic survey information for the period 1989-2006. Small clupeoids include sprat and 0-ring herring unless otherwise stated. CVs are approximate. Biomass in $t$. All surveys carried out at 38 kHz except December 1996, which was at 120 kHz .

| Year | Area | Dates | HERRING BIOMASS | CV | HERRING BIOMASS | CV | SMALL CLUPEOIDS | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 |  |  | (1+years) |  | (SSB) |  | biomass |  |
|  | Douglas Bank | 25-26 Sept |  |  | 18000 | - | - | - |
|  | Douglas Bank | 26-27 Sept |  |  | 26,600 | - | - | - |
| 1990 ( |  |  |  |  |  |  |  |  |
| 1991 | Western Irish Sea | $\begin{aligned} & 26 \text { July - } 8 \\ & \text { Aug } \end{aligned}$ | 12,760 | 0.23 |  |  | $66,000^{1}$ | 0.20 |
| 1992 | Western Irish Sea | 20-31 July | 17,490 | 0.19 |  |  | 43,200 | 0.25 |
|  | $\begin{aligned} & \text { + IOM east } \\ & \text { coast } \end{aligned}$ |  |  |  |  |  |  |  |
| 1994 | Area VIIa(N) | $\begin{aligned} & 28 \text { Aug - } 8 \\ & \text { Sep } \end{aligned}$ | 31,400 | 0.36 | 25,133 | - | 68,600 | 0.10 |
|  | Douglas Bank | 22-26 Sept |  |  | 28,200 | - | - | - |
| 1995 | Area VIIa(N) | 11-22 Sept | 38,400 | 0.29 | 20,167 | - | 348,600 | 0.13 |
|  | Douglas Bank | 10-11 Oct |  | - | 9,840 | - | - | - |
|  | Douglas Bank | 23-24 Oct |  |  | 1,750 | 0.51 | - | - |
| 1996 | Area VIIa(N) | 2-12 Sept | 24,500 | 0.25 | 21426 | 0.25 | $-{ }^{2}$ | - |
| 1997 | Area VIIa(N)reduced | 8-12 Sept | 20,100 | 0.28 | 10,702 | 0.35 | 46,600 | 0.20 |
| 1998 | Area VIIa(N) | 8-14 Sept | 14,500 | 0.20 | 9,157 | 0.18 | 228,000 | 0.11 |
| 1999 | Area VIIa(N) | 6-17 Sept | 31,600 | 0.59 | 21,040 | 0.75 | 272,200 | 0.10 |
| 2000 | Area VIIa(N) | 11-21 Sept | 40,200 | 0.26 | 33,144 | 0.32 | 234,700 | 0.11 |
| 2001 | Area VIIa(N) | 10-18 Sept | 35,400 | 0.40 | 13,647 | 0.42 | 299,700 | 0.08 |
| 2002 | Area VIIa(N) | 9-20 Sept | 41,400 | 0.56 | 25,102 | 0.83 | 413,900 | 0.09 |
| 2003 | Area VIIa(N) | 7-20 Sept | 49,500 | 0.22 | 24,390 | 0.24 | 265,900 | 0.10 |
| 2004 | Area VIIa(N) | $\begin{aligned} & \text { 6-10, 15/16, } \\ & \text { 28/29 Sept } \end{aligned}$ | 34,437 | 0.41 | 21,593 | 0.41 | 281,000 | 0.07 |
| 2005 | Area VIIa(N) | $\begin{aligned} & 29 \text { Aug - } 14 \\ & \text { Sept } \end{aligned}$ | 36,866 | 0.37 | 31,445 | 0.42 | 141,900 | 0.10 |
| 2006 | Area VIIa(N) | $\begin{aligned} & 30 \text { Aug - } 9 \\ & \text { Sept } \end{aligned}$ | 33,136 | 0.24 | 16,332 | 0.22 | 143,200 | 0.09 |

${ }^{1}$ sprat only; ${ }^{2}$ Data can be made available for the IoM waters only

Table 7.2.5 Irish Sea Herring Division VIIa(N). Age-disaggregated acoustic estimates of herring abundance from the Northern Ireland surveys in September (ACAGE).

| AGE (RINGS) | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1994 | 66.8 | 68.3 | 73.5 | 11.9 | 9.3 | 7.6 | 3.9 | 10.1 |
| 1995 | 319.1 | 82.3 | 11.9 | 29.2 | 4.6 | 3.5 | 4.9 | 6.9 |
| 1996 | 11.3 | 42.4 | 67.5 | 9 | 26.5 | 4.2 | 5.9 | 5.8 |
| 1997 | 134.1 | 50 | 14.8 | 11 | 7.8 | 4.6 | 0.6 | 1.9 |
| 1998 | 110.4 | 27.3 | 8.1 | 9.3 | 6.5 | 1.8 | 2.3 | 0.8 |
| 1999 | 157.8 | 77.7 | 34 | 5.1 | 10.3 | 13.5 | 1.6 | 6.3 |
| 2000 | 78.5 | 103.4 | 105.3 | 27.5 | 8.1 | 5.4 | 4.9 | 2.4 |
| 2001 | 387.6 | 93.4 | 10.1 | 17.5 | 7.7 | 1.4 | 0.6 | 2.2 |
| 2002 | 391 | 71.9 | 31.7 | 24.8 | 31.3 | 14.8 | 2.8 | 4.5 |
| 2003 | 349.2 | 220 | 32 | 4.7 | 3.9 | 4.1 | 1 | 0.9 |
| 2004 | 241 | 115.5 | 29.6 | 15.4 | 2.1 | 2.3 | 0.2 | 0.2 |
| 2005 | 94.3 | 109.9 | 97.1 | 17 | 8 | 0.8 | 0.6 | 5.8 |
| 2006 | 374.7 | 96.6 | 15.6 | 10.0 | 0.5 | 0.4 | 0.5 | 0.5 |

Table 7.2.6 Irish Sea Herring Division VIIa(N). Larval production (10 ${ }^{11}$ ) indices for the Manx component. Table amended with Douglas Bank time series removed, see 2005.

| Year |  | Northeast Irish Sea |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Isle of Man |  | Northern Ireland |  |  |  |

SE = Standard Error

* 2005 Index value amended

Table 7.2.7 Irish Sea herring Division VIIa(N). Northern Ireland groundfish survey indices for herring (Nos. per 3 miles).

## (a) 0-ring herring: October survey

|  | Western Irish Sea |  |  | EASTERN Irish Sea |  |  |  | Total Irish Sea |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey | Mean | N.obs | SE | Mean | N.obs. | SE | Mean | N. obs | SE |  |
| 1991 | 54 | 34 | 22 |  |  |  |  |  |  |  |
| 1992 | 210 | 31 | 99 | 240 | 8 | 149 | 177 | 46 | 68 |  |
| 1993 | 633 | 26 | 331 | 498 | 10 | 270 | 412 | 44 | 155 |  |
| 1994 | 548 | 26 | 159 | 8 | 7 | 5 | 194 | 41 | 55 |  |
| 1995 | 67 | 22 | 23 | 35 | 9 | 18 | 37 | 35 | 11 |  |
| 1996 | 90 | 26 | 58 | 131 | 9 | 79 | 117 | 42 | 50 |  |
| 1997 | 281 | 26 | 192 | 68 | 9 | 42 | 138 | 43 | 70 |  |
| 1998 | 980 | 26 | 417 | 12 | 9 | 10 | 347 | 43 | 144 |  |
| 1999 | 389 | 26 | 271 | 90 | 9 | 29 | 186 | 43 | 96 |  |
| 2000 | 202 | 24 | 144 | 367 | 9 | 190 | 212 | 38 | 89 |  |
| 2001 | 553 | 26 | 244 | 236 | 11 | 104 | 284 | 45 | 93 |  |
| 2002 | 132 | 26 | 84 | 18 | 11 | 10 | 63 | 45 | 31 |  |
| 2003 | 1203 | 26 | 855 | 75 | 11 | 47 | 446 | 45 | 296 |  |
| 2004 | 838 | 26 | 292 | 447 | 11 | 191 | 469 | 45 | 125 |  |
| 2005 | 1516 | 26 | 1036 | 256 | 11 | 152 | 627 | 45 | 363 |  |
| 2006 | 4677 | 26 | 2190 | 2140 | 11 | 829 | 2468 | 45 | 822 |  |

(b) 1-ring herring: March Surveys.

|  | Western Irish Sea |  |  |  | Eastern Irish Sea |  |  |  | Total Irish Sea |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey | Mean | N.obs | SE | Mean | N.obs. | SE | Mean | N.obs | SE |  |  |
| 1992 | 392 | 20 | 198 | 115 | 10 | 73 | 190 | 34 | 77 |  |  |
| 1993 | 1755 | 27 | 620 | 175 | 10 | 66 | 681 | 45 | 216 |  |  |
| 1994 | 2472 | 25 | 1852 | 106 | 9 | 51 | 923 | 39 | 641 |  |  |
| 1995 | 1299 | 26 | 679 | 73 | 8 | 32 | 480 | 42 | 235 |  |  |
| 1996 | 1055 | 22 | 638 | 285 | 9 | 164 | 487 | 39 | 230 |  |  |
| 1997 | 1473 | 26 | 382 | 260 | 9 | 96 | 612 | 43 | 137 |  |  |
| 1998 | 3953 | 26 | 1331 | 250 | 9 | 184 | 1472 | 43 | 466 |  |  |
| 1999 | 5845 | 26 | 1860 | 736 | 9 | 321 | 2308 | 42 | 655 |  |  |
| 2000 | 2303 | 26 | 853 | 546 | 10 | 217 | 1009 | 44 | 306 |  |  |
| 2001 | 3518 | 26 | 916 | 1265 | 11 | 531 | 1763 | 45 | 381 |  |  |
| $2002^{\text {a }}$ | 2255 | 25 | 845 | 185 | 11 | 84 | 852 | 44 | 294 |  |  |
| $2002^{\text {b }}$ | 7870 | 26 | 5667 | 185 | 11 | 84 | 2794 | 45 | 1960 |  |  |
| 2003 | 2103 | 26 | 876 | 896 | 11 | 604 | 1079 | 45 | 382 |  |  |
| 2004 | 6611 | 25 | 2726 | 491 | 11 | 163 | 2486 | 44 | 945 |  |  |
| 2005 | 7274 | 26 | 3097 | 1240 | 8 | 375 | 3001 | 42 | 1121 |  |  |
| 2006 | 4249 | 26 | 1687 | 2630 | 11 | 813 | 2496 | 45 | 662 |  |  |

a. Unusually large catch removed, $b$. unusually large catch retained.

Table 7.2.7 Continued. Irish Sea herring Division VIIa(N). Northern Ireland groundfish survey indices for herring (Nos. per 3 miles.).
(c) 1-ring herring: October Surveys

|  | Western Irish Sea |  |  |  | Eastern Irish Sea |  |  |  | Total Irish Sea |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: |
| Survey | Mean | N.obs | SE | Mean | N.obs. | SE | Mean | N.obs | SE |  |  |
| 1991 | 102 | 34 | 34 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |  |  |
| 1992 | 36 | 31 | 18 | 20 | 8 | 11 | 21 | 46 | 8 |  |  |
| 1993 | 122 | 26 | 66 | 4 | 10 | 2 | 44 | 44 | 23 |  |  |
| 1994 | 490 | 26 | 137 | 17 | 6 | 10 | 176 | 40 | 47 |  |  |
| 1995 | 153 | 22 | 61 | 3 | 9 | 1 | 55 | 35 | 21 |  |  |
| 1996 | 30 | 26 | 13 | 2 | 9 | 1 | 11 | 42 | 5 |  |  |
| 1997 | 612 | 26 | 369 | 0.2 | 9 | 0.2 | 302 | 43 | 156 |  |  |
| 1998 | 39 | 26 | 15 | 13 | 9 | 10 | 53 | 43 | 35 |  |  |
| 199 | 81 | 26 | 41 | 104 | 9 | 95 | 74 | 43 | 40 |  |  |
| 2000 | 455 | 24 | 250 | 74 | 9 | 52 | 579 | 38 | 403 |  |  |
| 2001 | 1412 | 26 | 641 | 5 | 11 | 3 | 513 | 45 | 223 |  |  |
| 2002 | 370 | 26 | 111 | 4 | 11 | 2 | 291 | 45 | 158 |  |  |
| 2003 | 314 | 26 | 143 | 410 | 11 | 350 | 267 | 45 | 144 |  |  |
| 2004 | 710 | 26 | 298 | 103 | 11 | 74 | 299 | 45 | 108 |  |  |
| 2005 | 3217 | 25 | 1467 | 18 | 11 | 12 | 1121 | 44 | 507 |  |  |
| 2006 | 1458 | 26 | 669 | 40 | 11 | 18 | 523 | 45 | 231 |  |  |

Table 7.3.1 Irish Sea Herring Division VIIa(N). Mean length-at-age in the catch.

| Year | Lengths-at-age (cm) <br> Age (rings) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1985 | 22.1 | 24.3 | 26.1 | 27.6 | 28.3 | 28.6 | 29.5 | 30.1 |
| 1986 | 19.7 | 24.3 | 25.8 | 26.9 | 28.0 | 28.8 | 28.8 | 29.8 |
| 1987 | 20.0 | 24.1 | 26.3 | 27.3 | 28.0 | 29.2 | 29.4 | 30.1 |
| 1988 | 20.2 | 23.5 | 25.7 | 26.3 | 27.2 | 27.7 | 28.7 | 29.6 |
| 1989 | 20.9 | 23.8 | 25.8 | 26.8 | 27.8 | 28.2 | 28.0 | 29.5 |
| 1990 | 20.1 | 24.2 | 25.6 | 26.2 | 27.7 | 28.3 | 28.3 | 29.0 |
| 1991 | 20.5 | 23.8 | 25.4 | 26.1 | 26.8 | 27.3 | 27.7 | 28.7 |
| 1992 | 19.0 | 23.7 | 25.3 | 26.2 | 26.7 | 27.2 | 27.9 | 29.4 |
| 1993 | 21.6 | 24.1 | 25.9 | 26.7 | 27.2 | 27.6 | 28.0 | 28.7 |
| 1994 | 20.1 | 23.9 | 25.5 | 26.5 | 27.0 | 27.4 | 27.9 | 28.4 |
| 1995 | 20.4 | 23.6 | 25.2 | 26.3 | 26.8 | 27.0 | 27.6 | 28.3 |
| 1996 | 19.8 | 23.5 | 25.3 | 26.0 | 26.6 | 27.6 | 27.6 | 28.2 |
| 1997 | 19.6 | 23.6 | 25.1 | 26.0 | 26.5 | 27.1 | 27.7 | 28.2 |
| 1998 | 20.8 | 23.8 | 25.2 | 26.1 | 27.0 | 26.8 | 27.2 | 28.7 |
| 1999 | 19.8 | 23.6 | 25.0 | 26.1 | 26.5 | 27.1 | 27.2 | 28.0 |
| 2000 | 19.7 | 23.8 | 25.3 | 26.3 | 27.1 | 27.7 | 27.7 | 28.1 |
| 2001 | 20.0 | 22.9 | 24.8 | 25.7 | 26.2 | 26.9 | 27.5 | 27.8 |
| 2002 | 21.1 | 23.1 | 24.8 | 26.0 | 26.6 | 26.7 | 27.0 | 28.1 |
| 2003 | 21.1 | 23.7 | 25.0 | 26.5 | 26.9 | 27.1 | 27.8 | 28.5 |
| 2004 | 20.7 | 23.1 | 24.6 | 25.8 | 26.1 | 27.1 | 27.6 | 28.3 |
| 2005 | 20.0 | 22.6 | 24.5 | 25.5 | 26.0 | 26.6 | 27.1 | 27.8 |
| 2006 | 19.5 | 22.7 | 24.3 | 25.3 | 26.0 | 26.6 | 26.9 | 28.0 |

Table 7.3.2 Irish Sea Herring Division VIIa(N). Mean weights-at-age in the catch.

| Year | Weights-at-age (g) <br> Age (rings) |  |  | 1 | 2 | 3 | 4 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

* Average for the preceding five years

Table 7.3.3 Irish Sea Herring Division VIIa(N). Maturity ogive (maturity in the catch).

| Year | Age (RINGS) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1961 | 0.00 | 0.22 | 0.63 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1962 | 0.00 | 0.24 | 0.83 | 0.92 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1963 | 0.00 | 0.34 | 0.88 | 0.89 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1964 | 0.00 | 0.53 | 0.81 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1965 | 0.00 | 0.61 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1966 | 0.00 | 0.47 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1967 | 0.02 | 0.37 | 0.75 | 0.83 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1968 | 0.00 | 0.88 | 0.94 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1969 | 0.00 | 0.71 | 0.92 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1970 | 0.02 | 0.92 | 0.94 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1971 | 0.15 | 0.87 | 0.97 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1972 | 0.11 | 0.88 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1973 | 0.12 | 0.77 | 0.89 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1974 | 0.36 | 0.99 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1975 | 0.40 | 0.99 | 1.00 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1976 | 0.07 | 0.96 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1977 | 0.03 | 0.92 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1978 | 0.04 | 0.81 | 0.88 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1979 | 0.00 | 0.84 | 0.81 | 0.78 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1980 | 0.20 | 0.88 | 0.95 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1981 | 0.19 | 0.89 | 0.90 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1982 | 0.10 | 0.80 | 0.89 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1983 | 0.02 | 0.73 | 0.88 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1984 | 0.00 | 0.69 | 0.83 | 0.93 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1985 | 0.14 | 0.62 | 0.71 | 0.88 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1986 | 0.31 | 0.73 | 0.66 | 0.81 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1987 | 0.00 | 0.85 | 0.91 | 0.87 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1988 | 0.00 | 0.90 | 0.96 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1989 | 0.07 | 0.63 | 0.93 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1990 | 0.06 | 0.66 | 0.90 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1991 | 0.04 | 0.30 | 0.74 | 0.82 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1992 | 0.28 | 0.48 | 0.72 | 0.81 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1993 | 0.00 | 0.46 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1994 | 0.19 | 0.68 | 0.99 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1995 | 0.10 | 0.86 | 0.94 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1996 | 0.02 | 0.60 | 0.96 | 0.83 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.04 | 0.82 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1998 | 0.30 | 0.83 | 0.97 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1999 | 0.02 | 0.84 | 0.95 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2000 | 0.14 | 0.79 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2001 | 0.15 | 0.54 | 0.88 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2002 | 0.02 | 0.92 | 0.95 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2003* | 0.11 | 0.76 | 0.95 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2004 | 0.11 | 1.00 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.20 | 0.97 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2006 | 0.19 | 0.89 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

[^11]Table 7.5.1 Irish Sea herring VIIa(N). ICA run $\log$ for the maximun-likelihood ICA calculation for the 6 year separable period. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

> Integrated Catch at Age Analysis

Version 1.4 w
K.R.Patterson

Fisheries Research Services
Marine Laboratory
Aberdeen
8 March 1998
Enter the name of the index file -->index.txt canum.txt
weca.txt
Stock weights in 2007 used for the year 2006
west.txt
Natural mortality in 2007 used for the year 2006
natmor.txt
Maturity ogive in 2007 used for the year 2006
matprop.txt
Name of age-structured index file (Enter if none) : -->fleet.txt
Name of the SSB index file (Enter if none) -->ssb.txt
No of years for separable constraint ?--> 6
Reference age for separable constraint ?--> 4
Constant selection pattern model (Y/N) ?-->y
S to be fixed on last age ?--> 1.000000000000000
First age for calculation of reference F ?--> 2
Last age for calculation of reference $F$ ?--> 6
Use default weighting (Y/N) ?-->n
Enter relative weights at age
Weight for age 1--> 0.100000000000000
Weight for age 2--> 1.000000000000000
Weight for age 3--> 1.000000000000000
Weight for age 4--> 1.000000000000000
Weight for age 5--> 1.000000000000000
Weight for age 6--> 1.000000000000000
Weight for age 7--> 1.000000000000000
Weight for age 8--> 1.000000000000000
Enter relative weights by year
Weight for year 2001--> 1.000000000000000
Weight for year 2002--> 1.000000000000000
Weight for year 2003--> 1.000000000000000
Weight for year 2004--> 1.000000000000000
Weight for year 2005--> 1.000000000000000
Weight for year 2006--> 1.000000000000000
Enter new weights for specified years and ages if needed
Enter year, age, new weight or $-1,-1,-1$ to end. $-1-1-1.000000000000000$
Is the last age of FLT01: Northern Ireland acoustic surveys a plus-group (Y-->y
You must choose a catchability model for each index.
Models: A Absolute: Index = Abundance . e
L Linear: Index $=\mathrm{Q}$. Abundance. e
P Power: Index = Q. Abundance^ K .e
where Q and K are parameters to be estimated, and
e is a lognormally-distributed error.

## Table 7.5.1 Irish Sea herring VIIa(N). ICA run log. Continued.

Model for NINEL is to be A/L/P ?-->L
Model for FLT01: Northern Ireland acoustic surveys is to be A/L/P ?-->L
Fit a stock-recruit relationship (Y/N) ?-->n
Enter lowest feasible F--> 5.0000000000000003E-02
Enter highest feasible F--> 2.000000000000000
Mapping the F-dimension of the SSQ surface

| F | SSQ |
| :---: | :---: |
| $+---------------------1998 ~$ |  |
| 0.05 | 40.5170011998 |
| 0.15 | 25.0575743055 |
| 0.26 | 21.7289232269 |
| 0.36 | 20.2684502580 |
| 0.46 | 19.4963702773 |
| 0.56 | 19.0839131174 |
| 0.67 | 18.9080410189 |
| 0.77 | 18.9640071024 |
| 0.87 | 19.0929695403 |
| 0.97 | 19.2200050477 |
| 1.08 | 19.3624089461 |
| 1.18 | 19.5210624269 |
| 1.28 | 19.6900076497 |
| 1.38 | 19.8652996585 |
| 1.49 | 20.0443517626 |
| 1.59 | 20.2255253074 |
| 1.69 | 20.4078796970 |
| 1.79 | 20.5910148832 |
| 1.90 | 20.7749871199 |
| 2.00 | 20.9602857842 |
| Lowest SSQ is for F $=$ | 0.698 |

No of years for separable analysis : 6
Age range in the analysis : $1 \ldots 8$
Year range in the analysis : $1961 \ldots 2006$
Number of indices of SSB : 1
Number of age-structured indices : 1
Parameters to estimate : 32
Number of observations : 160
Conventional selection $\quad$ single voctor to bel fitted.
Survey weighting to be Manual (recommended) or Iterative (M/I) ?-->M
Enter weight for NINEL--> 1.000000000000000
Enter weight for FLT01: Northern Ireland acoustic surveys at age 1--> 0.100000000000000
Enter weight for FLT01: Northern Ireland acoustic surveys at age 2--> 1.000000000000000
Enter weight for FLT01: Northern Ireland acoustic surveys at age 3--> 1.000000000000000
Enter weight for FLT01: Northern Ireland acoustic surveys at age 4--> 1.000000000000000
Enter weight for FLT01: Northern Ireland acoustic surveys at age 5--> 1.000000000000000
Enter weight for FLT01: Northern Ireland acoustic surveys at age 6--> 1.000000000000000
Enter weight for FLT01: Northern Ireland acoustic surveys at age 7--> 1.000000000000000
Enter weight for FLT01: Northern Ireland acoustic surveys at age 8--> 1.000000000000000
Enter estimates of the extent to which errors
in the age-structured indices are correlated
across ages. This can be in the range 0 (independence)
to 1 (correlated errors).
Enter value for FLT01: Northern Ireland acoustic surveys--> 1.000000000000000
Do you want to shrink the final fishing mortality (Y/N) ?-->N
Seeking solution. Please wait.
SSB index weights
1.000

Aged index weights
FLT01: Northern Ireland acoustic surveys
Age : $\begin{array}{llllllll}1 & 2 & 3 & 4 & 5 & 6 & 7 & 8\end{array}$
Wts : $\quad 0.0120 .1250 .1250 .1250 .1250 .1250 .1250 .125$
$F$ in 2006 at age 4 is 1.177082 in iteration 1
Detailed, Normal or Summary output (D/N/S)-->D
Output page width in characters (e.g. 80..132) ?--> 80
Estimate historical assessment uncertainty ?-->y
Sample from Covariances or Bayes MCMC (C/B) ?-->C
Use default percentiles ( $\mathrm{Y} / \mathrm{N}$ ) ?-->y
How many samples to take ?--> 100
Enter SSB reference level (e.g. MBAL, Bpa..) [t]--> $1.0000000000000000 \mathrm{E}+04$
Succesful exit from ICA

Table 7.5.2 Irish Sea herring VIIa(N). Catch number-at-age (millions). N.B. In this table "age" refers to number of rings (winter rings in the otolith).

Output Generated by ICA Version 1.4

| Herring Irish Sea |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catch in Number |  |  |  |  |  |  |  |  |
| AGE | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 |
| 1 | 4.54 | 0.38 | 4.84 | 1.51 | 0.85 | 0.94 | 4.44 | 1.02 |
| 2 | 11.47 | 12.30 | 9.44 | 18.10 | 27.08 | 15.05 | 40.92 | 30.18 |
| 3 | 2.63 | 7.34 | 2.34 | 4.35 | 8.18 | 15.64 | 5.60 | 13.46 |
| 4 | 12.43 | 1.81 | 2.89 | 0.71 | 0.99 | 2.00 | 4.63 | 4.08 |
| 5 | 0.24 | 5.43 | 2.26 | 0.53 | 0.71 | 0.12 | 1.35 | 0.82 |
| 6 | 0.48 | 0.19 | 2.26 | 0.71 | 0.99 | 0.35 | 0.00 | 0.61 |
| 7 | 1.20 | 0.19 | 0.55 | 0.00 | 0.42 | 0.12 | 0.00 | 0.00 |
| 8 | 2.15 | 0.67 | 0.62 | 0.18 | 0.71 | 0.00 | 0.00 | 0.00 |
| $\begin{array}{r} \times 10 \wedge 6 \\ \text { Catch in Number } \end{array}$ |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |
| AGE | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| 1 | 1.32 | 5.61 | 12.17 | 40.64 | 42.15 | 43.25 | 33.33 | 34.74 |
| 2 | 42.80 | 31.18 | 66.92 | 46.66 | 32.74 | 109.55 | 48.24 | 56.16 |
| 3 | 16.91 | 33.63 | 31.94 | 26.95 | 38.24 | 39.75 | 39.41 | 20.78 |
| 4 | 12.68 | 16.47 | 29.41 | 13.18 | 11.49 | 24.51 | 10.84 | 15.22 |
| 5 | 1.32 | 12.61 | 5.07 | 13.75 | 6.92 | 10.65 | 7.87 | 4.58 |
| 6 | 2.64 | 1.75 | 3.55 | 6.76 | 5.07 | 4.99 | 4.21 | 2.81 |
| 7 | 0.53 | 2.10 | 1.01 | 2.66 | 2.59 | 5.15 | 2.09 | 2.42 |
| 8 | 0.00 | 1.05 | 1.01 | 1.67 | 2.60 | 1.63 | 1.64 | 1.27 |


| AGE | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 30.28 | 15.54 | 11.77 | 5.84 | 5.05 | 5.10 | 1.31 | 1.17 |
| 2 | 39.04 | 36.95 | 38.27 | 25.76 | 15.79 | 16.03 | 12.16 | 8.42 |
| 3 | 22.69 | 13.41 | 23.49 | 19.51 | 3.20 | 5.67 | 5.60 | 7.24 |
| 4 | 6.75 | 6.78 | 4.25 | 8.52 | 2.79 | 2.15 | 2.82 | 3.84 |
| 5 | 4.52 | 1.74 | 2.20 | 1.98 | 2.30 | 0.33 | 0.45 | 2.22 |
| 6 | 1.46 | 1.34 | 1.05 | 0.91 | 0.33 | 1.11 | 0.48 | 0.38 |
| 7 | 0.91 | 0.67 | 0.40 | 0.36 | 0.29 | 0.14 | 0.26 | 0.23 |
| 8 | 1.12 | 0.35 | 0.29 | 0.23 | 0.24 | 0.38 | 0.06 | 0.48 |


| AGE | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2.43 | 4.49 | 2.23 | 2.61 | 1.16 | 2.31 | 2.00 | 12.15 |
| 2 | 10.05 | 15.27 | 12.98 | 21.25 | 6.39 | 12.84 | 9.75 | 6.89 |
| 3 | 17.34 | 7.46 | 6.15 | 13.34 | 12.04 | 5.73 | 6.74 | 6.74 |
| 4 | 13.29 | 8.55 | 3.00 | 7.16 | 4.71 | 9.70 | 2.83 | 6.69 |
| 5 | 7.21 | 4.53 | 4.18 | 4.61 | 1.88 | 3.60 | 5.07 | 3.26 |
| 6 | 2.65 | 3.20 | 2.78 | 5.08 | 1.25 | 1.66 | 1.49 | 5.12 |
| 7 | 0.67 | 1.46 | 2.33 | 3.23 | 1.56 | 1.04 | 0.72 | 1.04 |
| 8 | 0.72 | 0.88 | 1.67 | 4.21 | 1.96 | 1.62 | 0.81 | 0.39 |

Table 7.5.2 Irish Sea herring VIIa(N). Catch number-at-age (millions). Continued.

Catch in Number

| AGE | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.65 | 1.97 | 3.20 | 5.34 | 9.55 | 3.07 | 1.81 | 1.22 |
| 2 | 14.64 | 7.00 | 21.33 | 17.53 | 21.39 | 11.88 | 16.93 | 3.74 |
| 3 | 3.01 | 12.17 | 3.39 | 9.76 | 7.56 | 3.88 | 5.94 | 5.87 |
| 4 | 3.02 | 1.83 | 5.27 | 1.16 | 7.34 | 4.45 | 1.57 | 2.07 |
| 5 | 2.90 | 2.57 | 1.20 | 3.60 | 1.64 | 6.67 | 1.48 | 0.56 |
| 6 | 1.61 | 2.10 | 1.15 | 0.78 | 2.28 | 1.03 | 1.99 | 0.35 |
| 7 | 2.18 | 1.28 | 0.93 | 0.96 | 0.84 | 2.05 | 0.44 | 0.25 |
| 8 | 0.85 | 1.99 | 1.45 | 1.36 | 1.43 | 0.45 | 0.62 | 0.15 |


| AGE | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2.71 | 0.18 | 0.69 | 3.23 | 8.69 | 5.67 |
| 2 | 11.47 | 9.02 | 4.69 | 8.83 | 13.98 | 15.25 |
| 3 | 7.15 | 1.89 | 3.35 | 5.41 | 10.56 | 8.20 |
| 4 | 13.05 | 1.87 | 2.56 | 2.16 | 3.29 | 6.32 |
| 5 | 3.39 | 2.40 | 0.88 | 0.62 | 1.42 | 1.33 |
| 6 | 0.94 | 0.95 | 2.95 | 0.21 | 0.41 | 0.61 |
| 7 | 0.65 | 0.47 | 0.87 | 0.67 | 0.29 | 0.26 |
| 8 | 0.80 | 0.34 | 0.61 | 0.13 | 0.37 | 0.25 |

Table 7.5.3 Irish Sea herring VIIa(N). Weight in the catch (kg). N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| AGE | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.08200 | 0.06700 | 0.06700 | 0.07800 | 0.06500 | 0.09200 | 0.09300 | 0.09100 |
| 2 | 0.12300 | 0.12500 | 0.13100 | 0.12900 | 0.13200 | 0.14000 | 0.14900 | 0.15300 |
| 3 | 0.17800 | 0.15200 | 0.18400 | 0.15600 | 0.17600 | 0.18500 | 0.18000 | 0.19600 |
| 4 | 0.19800 | 0.17700 | 0.20800 | 0.17100 | 0.19200 | 0.21800 | 0.19900 | 0.23100 |
| 5 | 0.23200 | 0.19900 | 0.22800 | 0.22600 | 0.21000 | 0.25800 | 0.22300 | 0.24600 |
| 6 | 0.22600 | 0.21400 | 0.23400 | 0.24000 | 0.23000 | 0.25300 | 0.24300 | 0.26900 |
| 7 | 0.25300 | 0.27500 | 0.26600 | 0.00000 | 0.27200 | 0.22500 | 0.22700 | 0.23400 |
| 8 | 0.24800 | 0.25100 | 0.25800 | 0.29600 | 0.26500 | 0.26400 | 0.27500 | 0.26400 |

## Table 7.5.3 Irish Sea herring VIIa(N). Weight in the catch (kg). Continued.

| AGE | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.07400 | 0.10100 | 0.10800 | 0.07400 | 0.07400 | 0.07400 | 0.07400 | 0.07400 |
| 2 | 0.15200 | 0.16200 | 0.15800 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 |
| 3 | 0.20400 | 0.20600 | 0.18900 | 0.19500 | 0.19500 | 0.19500 | 0.19500 | 0.19500 |
| 4 | 0.23100 | 0.22500 | 0.21400 | 0.21900 | 0.21900 | 0.21900 | 0.21900 | 0.21900 |
| 5 | 0.25400 | 0.24500 | 0.22500 | 0.23200 | 0.23200 | 0.23200 | 0.23200 | 0.23200 |
| 6 | 0.26600 | 0.25100 | 0.26600 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 |
| 7 | 0.23900 | 0.26900 | 0.24100 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 |
| 8 | 0.27000 | 0.25800 | 0.24100 | 0.27800 | 0.27800 | 0.27800 | 0.27800 | 0.27800 |
| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |  |
| AGE | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 0.07400 | 0.07400 | 0.07400 | 0.07400 | 0.07400 | 0.07400 | 0.07400 | 0.07600 |
| 2 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.15500 | 0.14200 |
| 3 | 0.19500 | 0.19500 | 0.19500 | 0.19500 | 0.19500 | 0.19500 | 0.19500 | 0.18700 |
| 4 | 0.21900 | 0.21900 | 0.21900 | 0.21900 | 0.21900 | 0.21900 | 0.21900 | 0.21300 |
| 5 | 0.23200 | 0.23200 | 0.23200 | 0.23200 | 0.23200 | 0.23200 | 0.23200 | 0.22100 |
| 6 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.25100 | 0.24300 |
| 7 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.25800 | 0.24000 |
| 8 | 0.27800 | 0.27800 | 0.27800 | 0.27800 | 0.27800 | 0.27800 | 0.27800 | 0.27300 |
| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |  |
| AGE | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| 1 | 0.08700 | 0.06800 | 0.05800 | 0.07000 | 0.08100 | 0.09600 | 0.07300 | 0.06200 |
| 2 | 0.12500 | 0.14300 | 0.13000 | 0.12400 | 0.12800 | 0.14000 | 0.12300 | 0.11400 |
| 3 | 0.15700 | 0.16700 | 0.16000 | 0.16000 | 0.15500 | 0.16600 | 0.15500 | 0.14000 |
| 4 | 0.18600 | 0.18800 | 0.17500 | 0.17000 | 0.17400 | 0.17500 | 0.17100 | 0.15500 |
| 5 | 0.20200 | 0.21500 | 0.19400 | 0.18000 | 0.18400 | 0.18700 | 0.18100 | 0.16500 |
| 6 | 0.20900 | 0.22800 | 0.21000 | 0.19800 | 0.19500 | 0.19500 | 0.19000 | 0.17400 |
| 7 | 0.22200 | 0.23900 | 0.21800 | 0.21200 | 0.20500 | 0.20700 | 0.19800 | 0.18100 |
| 8 | 0.25800 | 0.25400 | 0.22900 | 0.23200 | 0.21800 | 0.21800 | 0.21700 | 0.19700 |
| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |  |
| AGE | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 1 | 0.08900 | 0.07000 | 0.07500 | 0.06700 | 0.06400 | 0.08000 | 0.06900 | 0.06400 |
| 2 | 0.12700 | 0.12300 | 0.12100 | 0.11600 | 0.11800 | 0.12300 | 0.12000 | 0.12000 |
| 3 | 0.15700 | 0.15300 | 0.14600 | 0.14800 | 0.14600 | 0.14800 | 0.14500 | 0.14800 |
| 4 | 0.17100 | 0.17000 | 0.16400 | 0.16200 | 0.16500 | 0.16300 | 0.16700 | 0.16800 |
| 5 | 0.18200 | 0.18000 | 0.17600 | 0.17700 | 0.17600 | 0.18100 | 0.17600 | 0.18800 |
| 6 | 0.19100 | 0.18900 | 0.18100 | 0.19900 | 0.18800 | 0.17700 | 0.18800 | 0.20400 |
| 7 | 0.19800 | 0.20200 | 0.19300 | 0.20000 | 0.20400 | 0.18800 | 0.19000 | 0.20000 |
| 8 | 0.21200 | 0.21200 | 0.20700 | 0.21400 | 0.21600 | 0.22200 | 0.21000 | 0.21300 |
| Weights at age in the catches ( Kg ) |  |  |  |  |  |  |  |  |
| AGE | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |  |  |
| 1 | $\begin{aligned} & 0.06700 \\ & 0.10600 \end{aligned}$ | 0.08500 0.08100 0.07300 0.06700 0.06400 |  |  |  |  |  |  |
| 2 |  | $0.11300$ | 0.11600 | 0.10700 | 0.10300 | 0.10500 |  |  |
| 3 | $\begin{aligned} & 0.10600 \\ & 0.13900 \end{aligned}$ | $0.14400$ | 0.13600 | 0.13000 | 0.13600 | 0.13100 |  |  |
| 4 | 0.15600 | 0.16700 | 0.16000 | 0.15700 | 0.15600 | 0.14900 |  |  |
| 5 | 0.16800 | 0.18000 | 0.16700 | 0.16500 | 0.16600 | 0.16400 |  |  |
| 6 | $0.18500$ | 0.18400 | 0.17200 | 0.18700 | 0.18000 | 0.17700 |  |  |
| 7 |  | $\begin{aligned} & 0.19100 \\ & 0.21700 \end{aligned}$ | 0.18600 | 0.20000 | 0.19100 | 0.18400 |  |  |
| 8 | 0.20500 |  | 0.19900 | 0.20500 | 0.20900 | 0.21000 |  |  |

Table 7.5.4 Irish Sea herring VIIa(N). Weight in the stock (kg). N.B. In this table "age" refers to number of rings (winter rings in the otolith).


Table 7.5.4 Irish Sea herring VIIa(N). Weight in the stock (kg). Continued. Weights at age in the stock ( Kg )

| AGE | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.06600 | 0.08500 | 0.08100 | 0.06700 | 0.06700 | 0.06400 |
| 2 | 0.10500 | 0.11300 | 0.11600 | 0.11400 | 0.10300 | 0.10500 |
| 3 | 0.13900 | 0.14400 | 0.13600 | 0.14400 | 0.13600 | 0.13100 |
| 4 | 0.15600 | 0.16700 | 0.16000 | 0.16100 | 0.15600 | 0.14900 |
| 5 | 0.16700 | 0.18000 | 0.16700 | 0.17000 | 0.16600 | 0.16400 |
| 6 | 0.18300 | 0.18400 | 0.17200 | 0.19200 | 0.18000 | 0.17700 |
| 7 | 0.19900 | 0.19100 | 0.18600 | 0.20200 | 0.19100 | 0.18400 |
| 8 | 0.20500 | 0.21700 | 0.19900 | 0.20500 | 0.20700 | 0.22000 |

Table 7.5.5 Irish Sea herring VIIa(N). Natural mortality. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| AGE | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |


| Natural Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |


| Natural Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |

Table 7.5.5 Irish Sea herring VIIa(N). Natural mortality. Continued.

| AGE | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |


| Natural Mortality (per year) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |


| Natural Mortality (per year) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 2 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 | 0.3000 |
| 3 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 | 0.2000 |
| 4 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 5 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 6 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 7 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 8 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.0000 |

Table 7.5.6 Irish Sea herring VIIa(N). Proportion mature. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

Proportion of fish spawning

| AGE | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0200 | 0.0000 |
| 2 | 0.2200 | 0.2400 | 0.3400 | 0.5300 | 0.6100 | 0.4700 | 0.3700 | 0.8800 |
| 3 | 0.6300 | 0.8300 | 0.8800 | 0.8100 | 0.9000 | 0.9100 | 0.7500 | 0.9400 |
| 4 | 1.0000 | 0.9200 | 0.8900 | 1.0000 | 1.0000 | 1.0000 | 0.8300 | 0.9400 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Table 7.5.6 Irish Sea herring VIIa(N). Proportion mature. Continued.
Proportion of fish spawning

| AGE | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0000 | 0.0200 | 0.1500 | 0.1100 | 0.1200 | 0.3600 | 0.4000 | 0.0700 |
| 2 | 0.7100 | 0.9200 | 0.8700 | 0.8800 | 0.7700 | 0.9900 | 0.9900 | 0.9600 |
| 3 | 0.9200 | 0.9400 | 0.9700 | 0.9000 | 0.8900 | 0.9600 | 1.0000 | 0.9800 |
| 4 | 0.9400 | 0.9600 | 0.9800 | 1.0000 | 0.9700 | 1.0000 | 0.9400 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |


| Proportion of fish spawning |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 0.0300 | 0.0400 | 0.0000 | 0.2000 | 0.1900 | 0.1000 | 0.0200 | 0.0000 |
| 2 | 0.9200 | 0.8100 | 0.8400 | 0.8800 | 0.8900 | 0.8000 | 0.7300 | 0.6900 |
| 3 | 0.9600 | 0.8800 | 0.8100 | 0.9500 | 0.9000 | 0.8900 | 0.8800 | 0.8300 |
| 4 | 1.0000 | 0.9100 | 0.7800 | 0.9500 | 0.9400 | 0.9100 | 0.9000 | 0.9300 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Proportion of fish spawning

| AGE | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.1400 | 0.3100 | 0.0000 | 0.0000 | 0.0700 | 0.0600 | 0.0400 | 0.2800 |
| 2 | 0.6200 | 0.7300 | 0.8500 | 0.9000 | 0.6300 | 0.6600 | 0.3000 | 0.4800 |
| 3 | 0.7100 | 0.6600 | 0.9100 | 0.9600 | 0.9300 | 0.9000 | 0.7400 | 0.7200 |
| 4 | 0.8800 | 0.8100 | 0.8700 | 0.9900 | 0.9500 | 0.9500 | 0.8200 | 0.8100 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| Proportion of fish spawning |  |  |  |  |  |  |  |  |
| AGE | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 1 | 0.0000 | 0.1900 | 0.1000 | 0.0200 | 0.0400 | 0.3000 | 0.0200 | 0.1400 |
| 2 | 0.4600 | 0.6800 | 0.8600 | 0.6000 | 0.8200 | 0.8300 | 0.8400 | 0.7900 |
| 3 | 0.9900 | 0.9900 | 0.9400 | 0.9600 | 0.9500 | 0.9700 | 0.9500 | 0.9900 |
| 4 | 1.0000 | 0.9700 | 0.9900 | 0.8300 | 1.0000 | 0.9900 | 0.9700 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Proportion of fish spawning

| AGE | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.1500 | 0.0200 | 0.1100 | 0.1140 | 0.2000 | 0.1900 |
| 2 | 0.5400 | 0.9200 | 0.7600 | 1.0000 | 0.9700 | 0.8900 |
| 3 | 0.8800 | 0.9500 | 0.9500 | 0.9700 | 0.9900 | 1.0000 |
| 4 | 0.9700 | 0.9800 | 0.9700 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Table 7.5.7 Irish Sea herring VIIa(N). Indices of spawning biomass.


Table 7.5.8 Irish Sea herring VIIa(N). Tuning indices. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| AGE-STRUCTURED INDICES |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT01: Northern Ireland acoustic surveys |  |  |  |  |  |  |  |  |
| AGE | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | 66.83 | 319.12 | 11.34 | 134.15 | 110.44 | 157.76 | 78.52 | 387.56 |
| 2 | 68.29 | 82.26 | 42.37 | 49.98 | 27.31 | 77.72 | 103.44 | 93.40 |
| 3 | 73.53 | 11.94 | 67.47 | 14.81 | 8.08 | 34.02 | 105.29 | 10.19 |
| 4 | 11.86 | 29.25 | 8.95 | 10.98 | 9.27 | 5.11 | 27.54 | 17.49 |
| 5 | 9.30 | 4.57 | 26.47 | 1.75 | 6.48 | 10.26 | 8.07 | 7.70 |
| 6 | 7.55 | 3.50 | 4.17 | 4.55 | 1.78 | 13.52 | 5.43 | 1.37 |
| 7 | 3.87 | 4.89 | 5.91 | 0.57 | 2.25 | 1.59 | 4.90 | 0.63 |
| 8 | 10.12 | 6.89 | 5.82 | 1.91 | 0.78 | 6.29 | 2.36 | 2.26 |
| $x 10 \wedge 3$ |  |  |  |  |  |  |  |  |
| FLT01: Northern Ireland acoustic surveys |  |  |  |  |  |  |  |  |
| AGE | 2002 | 2003 | 2004 | 2005 | 2006 |  |  |  |
| 1 | 390.98 | 349.22 | 241.01 | 94.33 | 374.73 |  |  |  |
| 2 | 71.94 | 220.01 | 115.53 | 109.94 | 96.62 |  |  |  |
| 3 | 31.70 | 31.98 | 29.59 | 97.11 | 15.63 |  |  |  |
| 4 | 24.80 | 4.74 | 15.40 | 17.02 | 9.98 |  |  |  |
| 5 | 31.28 | 3.92 | 2.07 | 8.03 | 0.53 |  |  |  |
| 6 | 14.83 | 4.09 | 2.30 | 0.81 | 0.37 |  |  |  |
| 7 | 2.76 | 0.98 | 0.24 | 0.61 | 0.48 |  |  |  |
| 8 | 4.46 | 0.91 | 0.24 | 5.80 | 0.05 |  |  |  |

Table 7.5.9 Irish Sea herring VIIa(N). Fishing mortality (per year). N.B. In this table "age" refers to number of rings (winter rings in the otolith).


Table 7.5.9 Irish Sea herring VIIa(N). Fishing mortality (per year). Continued.

| Fishing |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 0.0265 | 0.0155 | 0.0233 | 0.0154 | 0.0232 | 0.0299 |
| 2 | 0.5424 | 0.3165 | 0.4761 | 0.3157 | 0.4737 | 0.6102 |
| 3 | 0.7468 | 0.4358 | 0.6556 | 0.4347 | 0.6523 | 0.8402 |
| 4 | 1.0462 | 0.6105 | 0.9184 | 0.6089 | 0.9137 | 1.1771 |
| 5 | 0.8609 | 0.5024 | 0.7557 | 0.5011 | 0.7519 | 0.9686 |
| 6 | 0.7643 | 0.4460 | 0.6709 | 0.4448 | 0.6675 | 0.8599 |
| 7 | 1.0462 | 0.6105 | 0.9184 | 0.6089 | 0.9137 | 1.1771 |
| 8 | 1.0462 | 0.6105 | 0.9184 | 0.6089 | 0.9137 | 1.1771 |

Table 7.5.10 Irish Sea herring VIIa(N). Population abundance ( 1 January, millions). N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 |
| 1 | 64.45 | 52.49 | 125.15 | 216.52 | 118.79 | 359.36 | 345.12 | 553.89 |
| 2 | 32.38 | 21.09 | 19.09 | 43.24 | 78.78 | 43.21 | 131.65 | 124.38 |
| 3 | 10.45 | 14.28 | 5.35 | 6.21 | 16.76 | 35.43 | 19.27 | 62.82 |
| 4 | 23.28 | 6.20 | 5.15 | 2.29 | 1.24 | 6.43 | 15.03 | 10.75 |
| 5 | 1.45 | 9.32 | 3.89 | 1.93 | 1.40 | 0.20 | 3.92 | 9.21 |
| 6 | 0.85 | 1.08 | 3.31 | 1.39 | 1.24 | 0.60 | 0.07 | 2.27 |
| 7 | 2.38 | 0.32 | 0.80 | 0.86 | 0.58 | 0.20 | 0.21 | 0.05 |
| 8 | 4.28 | 1.10 | 0.91 | 0.57 | 0.97 | 0.35 | 0.19 | 0.25 |
| $\times 10 \wedge 6$ |  |  |  |  |  |  |  |  |
| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| AGE | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| 1 | 372.62 | 478.15 | 496.82 | 412.67 | 666.58 | 348.35 | 367.27 | 261.67 |
| 2 | 203.17 | 136.31 | 172.64 | 175.71 | 128.48 | 220.90 | 103.41 | 115.95 |
| 3 | 66.47 | 114.06 | 74.45 | 71.34 | 90.52 | 67.35 | 71.59 | 36.01 |
| 4 | 39.33 | 39.23 | 63.21 | 32.40 | 34.27 | 39.92 | 19.83 | 23.54 |
| 5 | 5.87 | 23.57 | 19.92 | 29.38 | 16.84 | 20.13 | 13.01 | 7.70 |
| 6 | 7.56 | 4.05 | 9.42 | 13.22 | 13.59 | 8.69 | 8.15 | 4.35 |
| 7 | 1.47 | 4.34 | 2.01 | 5.16 | 5.57 | 7.49 | 3.15 | 3.40 |
| 8 | 0.16 | 2.17 | 2.01 | 3.24 | 5.59 | 2.37 | 2.48 | 1.78 |
| $x 10 \wedge 6$ |  |  |  |  |  |  |  |  |
| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| AGE | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 319.58 | 243.50 | 133.68 | 142.89 | 195.72 | 213.34 | 215.55 | 122.54 |
| 2 | 76.42 | 100.17 | 80.61 | 42.41 | 49.19 | 69.07 | 75.53 | 78.54 |
| 3 | 38.68 | 23.84 | 42.95 | 27.52 | 9.94 | 23.05 | 37.53 | 45.58 |
| 4 | 11.01 | 11.51 | 7.59 | 14.26 | 5.31 | 5.27 | 13.78 | 25.69 |
| 5 | 6.96 | 3.60 | 4.02 | 2.86 | 4.86 | 2.17 | 2.73 | 9.79 |
| 6 | 2.65 | 2.04 | 1.61 | 1.56 | 0.72 | 2.23 | 1.65 | 2.05 |
| 7 | 1.29 | 1.02 | 0.58 | 0.47 | 0.55 | 0.34 | 0.97 | 1.03 |
| 8 | 1.58 | 0.53 | 0.42 | 0.30 | 0.45 | 0.92 | 0.22 | 2.16 |

Table 7.5.10 Irish Sea herring VIIa(N). Population abundance ( 1 January, millions). Continued.

| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| 1 | 140.44 | 163.37 | 253.28 | 102.90 | 141.03 | 108.69 | 64.73 | 191.15 |
| 2 | 44.40 | 50.26 | 57.50 | 91.88 | 36.34 | 51.21 | 38.64 | 22.65 |
| 3 | 50.98 | 24.34 | 24.28 | 31.55 | 49.99 | 21.48 | 27.02 | 20.33 |
| 4 | 30.80 | 26.20 | 13.23 | 14.36 | 13.90 | 30.11 | 12.44 | 16.07 |
| 5 | 19.59 | 15.30 | 15.61 | 9.13 | 6.22 | 8.11 | 18.05 | 8.57 |
| 6 | 6.75 | 10.91 | 9.55 | 10.16 | 3.90 | 3.85 | 3.94 | 11.53 |
| 7 | 1.49 | 3.60 | 6.84 | 6.01 | 4.39 | 2.34 | 1.91 | 2.15 |
| 8 | 1.62 | 2.16 | 4.91 | 7.83 | 5.50 | 3.63 | 2.17 | 0.81 |


| Population Abundance (1 January) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 1 | 61.48 | 194.79 | 126.35 | 112.38 | 122.56 | 195.48 | 56.96 | 60.15 |
| 2 | 63.31 | 22.24 | 70.51 | 44.62 | 38.26 | 39.59 | 70.13 | 19.90 |
| 3 | 10.94 | 34.45 | 10.54 | 34.14 | 18.25 | 10.44 | 19.25 | 37.55 |
| 4 | 10.60 | 6.26 | 17.30 | 5.59 | 19.19 | 8.18 | 5.08 | 10.43 |
| 5 | 8.21 | 6.73 | 3.93 | 10.66 | 3.95 | 10.41 | 3.19 | 3.11 |
| 6 | 4.67 | 4.68 | 3.66 | 2.42 | 6.23 | 2.03 | 3.13 | 1.49 |
| 7 | 5.59 | 2.71 | 2.24 | 2.22 | 1.45 | 3.48 | 0.86 | 0.96 |
| 8 | 2.17 | 4.22 | 3.51 | 3.15 | 2.47 | 0.77 | 1.20 | 0.56 |


| Population Abundance (1 January) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 1 | 72.71 | 67.15 | 122.33 | 116.51 | 134.03 | 936.17 | 131.70 |
| 2 | 21.42 | 26.05 | 24.32 | 43.97 | 42.20 | 48.18 | 334.26 |
| 3 | 11.56 | 9.23 | 14.06 | 11.19 | 23.75 | 19.47 | 19.39 |
| 4 | 25.46 | 4.48 | 4.88 | 5.98 | 5.93 | 10.13 | 6.88 |
| 5 | 7.48 | 8.09 | 2.20 | 1.76 | 2.94 | 2.15 | 2.82 |
| 6 | 2.29 | 2.86 | 4.43 | 0.94 | 0.97 | 1.25 | 0.74 |
| 7 | 1.02 | 0.96 | 1.66 | 2.05 | 0.54 | 0.45 | 0.48 |
| 8 | 1.29 | 0.77 | 1.05 | 0.29 | 0.64 | 0.36 | 0.23 |

Table 7.5.11 Irish Sea herring VIIa(N). Weighting factors in number. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| AGE | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 | 0.1000 |
| 2 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 3 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 6 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Table 7.5.12 Irish Sea herring VIIa(N). Predicted SSB Index values. N.B. In this table "age" refers to number of rings (winter rings in the otolith).


Table 7.5.13 Irish Sea herring VIIa(N). Predicted age-structured Index values. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

Predicted Age-Structured Index Values

| FLT01: Northern Ireland acoustic survey Predicted |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | 236.7 | 150.7 | 130.4 | 136.7 | 236.0 | 67.4 | 72.2 | 87.7 |
| 2 | 49.7 | 160.3 | 89.4 | 56.6 | 90.3 | 171.9 | 51.8 | 44.6 |
| 3 | 59.6 | 19.0 | 64.2 | 29.0 | 17.6 | 35.2 | 81.3 | 16.5 |
| 4 | 11.7 | 31.8 | 11.4 | 32.0 | 10.7 | 9.3 | 21.5 | 28.5 |
| 5 | 10.0 | 6.4 | 16.7 | 5.6 | 9.9 | 4.2 | 5.8 | 8.5 |
| 6 | 6.2 | 5.7 | 3.8 | 9.2 | 2.4 | 2.9 | 2.6 | 2.7 |
| 7 | 3.2 | 2.8 | 2.7 | 1.4 | 3.3 | 0.9 | 1.5 | 0.9 |
| 8 | 7.8 | 7.1 | 6.2 | 3.8 | 1.2 | 2.1 | 1.4 | 1.8 |



Table 7.5.14 Irish Sea herring VIIa(N). Fitted selection pattern. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| AGE | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.1436 | 0.0316 | 0.0714 | 0.0282 | 0.0066 | 0.0105 | 0.0527 | 0.0058 |
| 2 | 0.6367 | 2.9310 | 0.9360 | 1.6499 | 0.2894 | 1.2874 | 1.1284 | 0.6454 |
| 3 | 0.3963 | 2.2437 | 0.7378 | 3.5842 | 0.4399 | 1.6668 | 0.9838 | 0.5303 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.2334 | 2.5598 | 1.0605 | 0.8666 | 0.4328 | 2.4084 | 1.1483 | 0.1930 |
| 6 | 1.0908 | 0.5597 | 1.4122 | 1.9518 | 1.0000 | 2.4084 | 0.7305 | 0.6574 |
| 7 | 0.9121 | 2.7348 | 1.4122 | 1.0000 | 0.8138 | 2.4084 | 1.0000 | 1.0000 |
| 8 | 0.9121 | 2.7348 | 1.4122 | 1.0000 | 0.8138 | 2.4084 | 1.0000 | 1.0000 |
| Fitted Selection Pattern |  |  |  |  |  |  |  |  |
| AGE | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| 1 | 0.0136 | 0.0323 | 0.0591 | 0.3011 | 0.2416 | 0.2100 | 0.1809 | 0.2064 |
| 2 | 0.6730 | 0.5275 | 0.8767 | 0.6554 | 0.8000 | 0.8095 | 0.8932 | 0.7134 |
| 3 | 0.7944 | 0.6756 | 0.9490 | 0.9617 | 1.4310 | 1.0016 | 1.0791 | 0.8808 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.6539 | 1.4147 | 0.4660 | 1.2113 | 1.2991 | 0.7872 | 1.1781 | 0.8647 |
| 6 | 1.1057 | 1.0406 | 0.7531 | 1.3782 | 1.1453 | 0.8945 | 0.9169 | 0.9994 |
| 7 | 1.1439 | 1.2215 | 1.1236 | 1.3941 | 1.5389 | 1.2291 | 1.3843 | 1.2058 |
| 8 | 1.1439 | 1.2215 | 1.1236 | 1.3941 | 1.5389 | 1.2291 | 1.3843 | 1.2058 |
| Fitted Selection Pattern |  |  |  |  |  |  |  |  |
| AGE | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| 1 | 0.1571 | 0.1107 | 0.1687 | 0.0681 | 0.0522 | 0.0690 | 0.0398 | 0.0889 |
| 2 | 0.8489 | 0.5739 | 0.8829 | 1.1802 | 0.5759 | 0.5566 | 0.8489 | 0.7739 |
| 3 | 0.9936 | 0.9912 | 1.0288 | 1.4818 | 0.5471 | 0.5650 | 0.7418 | 1.1245 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.1062 | 0.7391 | 0.9665 | 1.3089 | 0.8569 | 0.3122 | 0.7762 | 1.5911 |
| 6 | 0.8394 | 1.2075 | 1.2969 | 0.9652 | 0.8186 | 1.3199 | 1.5194 | 1.2681 |
| 7 | 1.3046 | 1.2094 | 1.4189 | 1.6529 | 1.0079 | 1.0103 | 1.3400 | 1.5475 |
| 8 | 1.3046 | 1.2094 | 1.4189 | 1.6529 | 1.0079 | 1.0103 | 1.3400 | 1.5475 |
| Fitted Selection Pattern |  |  |  |  |  |  |  |  |
| AGE | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| 1 | 0.0461 | 0.1059 | 0.0515 | 0.0554 | 0.0298 | 0.0830 | 0.1829 | 0.1836 |
| 2 | 0.5020 | 1.0225 | 1.1071 | 0.4196 | 0.5157 | 0.8244 | 1.2542 | 0.7483 |
| 3 | 0.7763 | 0.9792 | 1.1999 | 0.8423 | 0.7009 | 0.8408 | 1.1732 | 0.7893 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.8101 | 0.8878 | 1.2148 | 1.0187 | 0.8663 | 1.5137 | 1.2773 | 0.8861 |
| 6 | 0.8817 | 0.8777 | 1.3400 | 1.0055 | 0.9373 | 1.4564 | 1.8527 | 1.0920 |
| 7 | 1.0495 | 1.3249 | 1.6254 | 1.1208 | 1.0625 | 1.5203 | 1.8303 | 1.2242 |
| 8 | 1.0495 | 1.3249 | 1.6254 | 1.1208 | 1.0625 | 1.5203 | 1.8303 | 1.2242 |
| Fitted Selection Pattern |  |  |  |  |  |  |  |  |
| AGE | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 1 | 0.0473 | 0.0441 | 0.1062 | 0.3158 | 0.2544 | 0.0299 | 0.1317 | 0.1399 |
| 2 | 0.8716 | 1.2244 | 1.1072 | 2.4201 | 1.9525 | 0.5017 | 0.8322 | 1.0478 |
| 3 | 1.0137 | 1.3386 | 1.1313 | 1.5317 | 1.1789 | 0.6203 | 1.0574 | 0.8113 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 1.3062 | 1.3943 | 1.0024 | 1.7781 | 1.1131 | 1.3119 | 1.6922 | 0.8957 |
| 6 | 1.2597 | 1.7412 | 1.0427 | 1.6776 | 0.9441 | 0.9023 | 2.7774 | 1.1992 |
| 7 | 1.4826 | 1.8630 | 1.4733 | 2.4543 | 1.8126 | 1.1332 | 1.9890 | 1.3776 |
| 8 | 1.4826 | 1.8630 | 1.4733 | 2.4543 | 1.8126 | 1.1332 | 1.9890 | 1.3776 |

Table 7.5.14 Irish Sea herring VIIa(N). Fitted selection pattern. Continued.

| AGE | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0254 | 0.0254 | 0.0254 | 0.0254 | 0.0254 | 0.0254 |
| 2 | 0.5184 | 0.5184 | 0.5184 | 0.5184 | 0.5184 | 0.5184 |
| 3 | 0.7138 | 0.7138 | 0.7138 | 0.7138 | 0.7138 | 0.7138 |
| 4 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 5 | 0.8229 | 0.8229 | 0.8229 | 0.8229 | 0.8229 | 0.8229 |
| 6 | 0.7305 | 0.7305 | 0.7305 | 0.7305 | 0.7305 | 0.7305 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 | 1.0000 |

Table 7.5.15 Irish Sea herring VIIa(N). Stock summary. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

STOCK SUMMARY


No of years for separable analysis : 6
Age range in the analysis : 1 . . . 8
Year range in the analysis : 1961 . . . 2006
Number of indices of SSB : 1
Number of age-structured indices : 1
Parameters to estimate : 32
Number of observations : 160
Conventional single selection vector model to be fitted.

Table 7.5.16 Irish Sea herring VIIa(N). Parameter estimates. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

PARAMETER ESTIMATES


Age-structured index catchabilities
FLT01: Northern Ireland acoustic survey

| Linear model fitted. Slopes at age : |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 25 | 1 | Q | 2.604 | 96 | 1.030 | 45.49 | 2.604 | 17.99 |
| 26 | 2 | Q | 3.916 | 31 | 2.905 | 9.836 | 3.916 | 7.296 |
| 27 | 3 | Q | 2.898 | 31 | 2.151 | 7.267 | 2.898 | 5.393 |
| 28 | 4 | Q | 2.641 | 31 | 1.958 | 6.636 | 2.641 | 4.922 |
| 29 | 5 | Q | 2.342 | 31 | 1.734 | 5.911 | 2.342 | 4.378 |
| 30 | 6 | Q | 2.284 | 31 | 1.686 | 5.824 | 2.284 | 4.299 |
| 31 | 7 | Q | 2.090 | 32 | 1.531 | 5.457 | 2.090 | 3.997 |
| 32 | 8 | $Q$ | 3.322 | 31 | 2.450 | 8.497 | 3.322 | 6.266 |

Table 7.5.17 Irish Sea herring VIIa(N). Residuals about the model fit. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

RESIDUALS ABOUT THE MODEL FIT

| Separable Model Residuals |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 0.810 | -1.295 | -0.944 | 1.048 | 1.497 | -1.125 |
| 2 | 0.379 | 0.383 | -0.540 | -0.160 | 0.004 | -0.235 |
| 3 | 0.248 | -0.452 | -0.616 | 0.406 | 0.012 | -0.215 |
| 4 | -0.194 | -0.049 | -0.095 | -0.188 | -0.036 | -0.064 |
| 5 | -0.201 | -0.243 | -0.238 | -0.064 | -0.046 | 0.034 |
| 6 | -0.223 | -0.031 | 0.352 | -0.410 | -0.083 | -0.137 |
| 7 | 0.020 | 0.119 | -0.091 | -0.284 | -0.066 | -0.130 |

SPAWNING BIOMASS INDEX RESIDUALS

| NINEL |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | *** | ***** | * | * | 0.384 | 0.879 | -0.778 | -1.748 |
| NINEL |  |  |  |  |  |  |  |  |
|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 1 | 0.258 | -1.479 | -0.476 | 0.195 | 1.335 | 0.687 | 0.157 | 0.105 |
| NINEL |  |  |  |  |  |  |  |  |
|  | 2005 | 2006 |  |  |  |  |  |  |
| 1 | 0.264 | 0.220 |  |  |  |  |  |  |
| AGE-STRUCTURED INDEX RESIDUALS |  |  |  |  |  |  |  |  |
| FLT01: Northern Ireland acoustic survey |  |  |  |  |  |  |  |  |
| Age | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | -1.265 | 0.750 | -2.442 | -0.019 | -0.759 | 0.850 | 0.084 | 1.486 |
| 2 | 0.317 | -0.667 | -0.746 | -0.124 | -1.195 | -0.794 | 0.691 | 0.739 |
| 3 | 0.211 | -0.463 | 0.049 | -0.670 | -0.779 | -0.035 | 0.259 | -0.479 |
| 4 | 0.017 | -0.083 | -0.240 | -1.070 | -0.141 | -0.598 | 0.249 | -0.487 |
| 5 | -0.071 | -0.335 | 0.461 | -1.164 | -0.424 | 0.886 | 0.334 | -0.101 |
| 6 | 0.205 | -0.496 | 0.103 | -0.703 | -0.313 | 1.524 | 0.749 | -0.688 |
| 7 | 0.205 | 0.542 | 0.769 | -0.899 | -0.383 | 0.533 | 1.209 | -0.368 |
| 8 | 0.260 | -0.027 | -0.061 | -0.688 | -0.394 | 1.110 | 0.550 | 0.221 |

FLT01: Northern Ireland acoustic survey

| Age | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1.566 | 0.859 | 0.531 | -0.541 | -1.100 |
| 2 | 0.113 | 1.419 | 0.063 | 0.173 | 0.013 |
| 3 | 0.647 | 0.400 | 0.384 | 0.983 | -0.504 |
| 4 | 1.273 | -0.238 | 0.507 | 0.843 | -0.028 |
| 5 | 0.953 | 0.367 | -0.242 | 0.792 | -1.451 |
| 6 | 1.229 | -0.328 | 0.481 | -0.428 | -1.330 |
| 7 | 0.847 | -0.502 | -2.358 | 0.135 | 0.283 |
| 8 | 1.088 | -0.584 | -0.862 | 1.763 | -2.363 |

Table 7.5.18 Irish Sea herring VIIa(N). Residuals about the model fit. N.B. In this table "age" refers to number of rings (winter rings in the otolith).
PARAMETERS OF THE DISTRIBUTION OF $\ln$ (CATCHES AT AGE)

| Separable model fitted from 2001 to 2006 |  |
| :--- | ---: |
| Variance | 0.1604 |
| Skewness test stat. | -1.3362 |
| Kurtosis test statistic | -0.7806 |
| Partial chi-square | 0.3876 |
| Sigificance in fit | 0.000 |
| Degrees of freedom | 19 |

## PARAMETERS OF DISTRIBUTIONS OF THE SSB INDICES

DISTRIBUTION STATISTICS FOR NINEL
Linear catchability relationship assumed
Last age is a plus-group

| Variance | 0.7314 |
| :--- | ---: |
| Skewness test stat. | -1.1024 |
| Kurtosis test statistic | -0.0949 |
| Partial chi-square | 3.0379 |
| Significance in fit | 0.0022 |
| Number of observations | 14 |
| Degrees of freedom | 13 |
| Weight in the analysis | 1.0000 |

## PARAMETERS OF THE DISTRIBUTION OF THE AGE-STRUCTURED INDICES

> DISTRIBUTION STATISTICS FOR FLT01: Northern Ireland acoustic survey

Linear catchability relationship assumed

|  |  |  |  |  |  |  |  |  |
| :--- | ---: | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| Var. | 0.0173 | 0.0652 | 0.0371 | 0.0472 | 0.0688 | 0.0838 | 0.1083 | 0.1380 |
| Skewness test stat. | -0.7499 | 0.2397 | 0.2030 | 0.6711 | -0.7876 | 0.6155 | -1.7267 | -0.6703 |
| Kurtosis test statisti | -0.3934 | -0.3673 | -0.7772 | -0.0180 | -0.4037 | -0.4711 | 0.8804 | 0.2343 |
| Partial chi-square | 0.0175 | 0.0698 | 0.0441 | 0.0607 | 0.0987 | 0.1272 | 0.1714 | 0.2468 |
| Significance in fit | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| Number of observations | 13 | 13 | 13 | 13 | 13 | 13 | 13 | 13 |
| Degrees of freedom | 12 | 12 | 12 | 12 | 12 | 12 | 12 | 12 |
| Weight in the analysis | 0.0125 | 0.1250 | 0.1250 | 0.1250 | 0.1250 | 0.1250 | 0.1250 | 0.1250 |

Table 7.5.19 Irish Sea herring VIIa(N). Analyses of variance. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| ANALYSIS OF VARIANCE |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Unweighted Statistics |  |  |  |  |  |
| Variance |  |  |  |  |  |
|  | SSQ | Data | Parameters | d.f. | Variance |
| Total for model | 88.8698 | 160 | 32 | 128 | 0.6943 |
| Catches at age | 10.0940 | 42 | 23 | 19 | 0.5313 |
| SSB Indices |  |  |  |  |  |
| NINEL | 9.5084 | 14 | 1 | 13 | 0.7314 |
| Aged Indices |  |  |  |  |  |
| FLT01: Northern Ireland acoustic surve Weighted Statistics | 69.2674 | 104 | 8 | 96 | 0.7215 |
| Variance |  |  |  |  |  |
|  | SSQ | Data | Parameters | d.f. | Variance |
| Total for model | 13.3817 | 160 | 32 | 128 | 0.1045 |
| Catches at age | 3.0481 | 42 | 23 | 19 | 0.1604 |
| SSB Indices |  |  |  |  |  |
| NINEL | 9.5084 | 14 | 1 | 13 | 0.7314 |
| Aged Indices |  |  |  |  |  |
| FLT01: Northern Ireland acoustic surve | 0.8252 | 104 | 8 | 96 | 0.0086 |



Figure 7.1.1
Irish Sea herring VIIa(N). Landings of herring from VIIa(N) from 1961 to 2006.


Figure 7.2.1 Irish Sea herring VIIa(N). Landings (catch-at-age) of herring from VIIa(N) from 1961 to 2006.


Figure 7.2.2 Irish Sea herring VIIa(N). A) Transects, stratum boundaries and trawl positions for the 2006 acoustic survey; (B) Density distribution of sprats (size of ellipses is proportional to square root of the fish density (t n.mile ${ }^{-2}$ ) per 15-minute interval). Maximum density was 330 t n.mile ${ }^{-2}$.


Figure 7.2.3
Irish Sea herring VIIa(N). (A) Density distribution of 1-ring and older herring (size of ellipses is proportional to square root of the fish density ( $\mathbf{t}$ n.mile ${ }^{-2}$ ) per $\mathbf{1 5}$-minute interval). Maximum density was 1100 t n.mile ${ }^{-2}$. (B) Density distribution of 0 -ring herring. Maximum density was 100 t n.mile ${ }^{-2}$. Note: same scaling of ellipse sizes on above figures.


Figure 7.2.4 Irish Sea herring VIIa(N). Percentage length compositions of herring in each trawl sample in the September 2006 acoustic survey. Trawl 1 and 2 excluded for representation purposes only.


Figure 7.2.5 Irish Sea herring VIIa(N). Estimates of larval herring abundance in the Northern Irish Sea, $6^{\text {th }}$ to $10^{\text {th }}$ November 2006. Areas of the circles are proportional to herring abundance (maximum abundance $=188$ per $\mathbf{m}^{2}$ ).


Figure 7.5.1 Irish Sea herring VIIa(N). SSQ surface for the deterministic calculation of the 6-year separable period.


Figure 7.5.2 Irish Sea herring VIIa(N). Comparison of mean reference $F_{2-6}$ for NINEL tuning index (NINEL), ACAGE tuning index (ACOU) and SPALY run (SPAL).


Figure 7.5.3 Irish Sea herring VIIa(N). Change in mean log abundance of year classes (1975 to 2002) per age class 2 to 7 (rings) for 5 year periods.


Figure 7.5.4 Irish Sea herring VIIa(N). Change in mean total mortality (Z) of year classes (1975 to 2002) for age classes (rings) 2 to 7 for 5 year periods.








Figure 7.5.5 Irish Sea herring VIIa(N). Inter-annual comparisons of proportion (\%) catch at age (rings) between catch and the acoustic survey.


Figure 7.5.6 Irish Sea herring VIIa(N). Time-series of shifts in adult herring (1+ring) biomass distribution between the west and east Isle of Man coasts estimated from acoustic surveys 1994-2006.


Figure 7.5.7 Irish Sea herring VIIa(N). Relationship between adult biomass distribution and ratio of age 2 proportion catch between acoustic and catch data. Year 2003 highlighted as year with largest discrepancy between catch and acoustic data.


Figure 7.5.8 Irish Sea herring VIIa(N). Illustration of selection patterns diagnostics, from the deterministic calculation (6-year separable period). Top left, a contour plot of selection pattern residuals. Top right, estimated selection (relative to age 4), error bars standard deviation. Bottom, marginal totals of residuals by year and age.


Figure 7.5.9 Irish Sea herring VIIa(N). Bubble plot of age and year residuals from ICA SPALY run (6-year separable period). Largest residual = -0.616. Black bubbles negative.


Figure 7.5.10 Irish Sea herring VIIa(N). Illustration of stock trends from deterministic calculation (6-year separable period). Summary of estimates of landings, fishing mortality-at-age 4, recruitment at age 1 and SSB at spawning.




Figure 7.5.11 Irish Sea herring VIIa(N). Retrospective trends in fishing mortality (F 2-6), SSB and recruitment (1-ringers) from ICA SPALY run.

## 8 Sprat in the North Sea

### 8.1 The Fishery

### 8.1.1 ACFM advice applicable for 2005 and 2006

From 2002 to 2005 the TAC set by management for Subarea IV (EU zone) and Division IIa (EU zone) has been 257000 t . The advice from 2005 was that a constant exploitation rate was expected to generate landings of 244000 in 2005.

ACFM in 2006 advised to set the TAC in 2006 to well below the TAC of 2005. Though relative trends in the biomass indicated an increase over most of the time series, the estimated recruitment of the 2005 year class to the 2006 fishery was low. The TAC set for 2006 was 282 700 t . A mid year revision of the TAC resulted in a final TAC for 2006 of 175000 t .

There have been no explicit management objectives for this stock.
For 2007 a preliminary TAC is set at 175000 t .

### 8.1.2 Total landings in 2006

Landing statistics for sprat for the North Sea by area and country are presented in Table 8.1.1 for 1996-2006. Landings data prior to 1996 are considered unreliable. As in previous years, sprat from the fjords of western Norway are not included in the landings for the North Sea. Landings from the fjords are presented separately (Table 8.1.2) due to uncertainties in stock identity. Table 8.1.3 shows the landings for 1996-2005 by year, quarter, and area in the North Sea. Generally, most of the landings are taken in the second half of the year. This was also seen in 2006, but about $41 \%$ were taken in the first quarter. The Norwegian vessels are not allowed to fish in the $2^{\text {nd }}$ and $3^{\text {rd }}$ quarters in the EU and the Norwegian zone and not allowed to fish in the Norwegian zone until the quota in the EU-zone has been taken.

The landings in 2006 were 113710 t, mainly taken by the Danish fleets. This was the lowest landings since 1997 and represents a reduction of $45 \%$ compared to 2005 where the landings were the highest recorded since mid 1990s. The Norwegian sprat fishery caught nearly 10000 tonnes of sprat in the North Sea in 2006 after three years with nearly no landings.

The quarterly and annual distributions of landings by rectangle for Subarea IV are shown in Figures 8.1.1-8.1.2.

### 8.2 Biological composition of the catch

### 8.2.1 By-catch in the North Sea sprat fishery

Only data on by-catch from the Danish fishery were available to the Working Group (Table 8.2.1). The Danish sprat fishery has in general been conducted with minor by-catch of herring. The total amount of herring caught as by-catch in the sprat fishery in 2006 is less than $10 \%$ of the total landings. The herring by-catch decreased in 2006 and was the lowest since 1999.

### 8.2.2 Landings in number

The biological sampling from 1996 onwards is considered reliable and the estimated quarterly landings-at-age in numbers for the period are presented in Table 8.2.2. Denmark and Norway provided age data of commercial landings in 2006 for all quarters fished. These data were used to raise the landings data from the North Sea. The landings by UK (England) were minor and unsampled (Table 8.2.3). In 1996-2005 1-ringer sprat dominated the landings (54-96\%),
while the 2-ringers in 2006 (2004 year class) made a larger part of the landings (47\%). The majority of the total sprat landings by numbers are normally taken in the second part of the year, but in 2006 more than $50 \%$ were taken in the first quarter.

### 8.2.3 Quality of landings and biological data

The sampling intensity for biological samples, i.e., age and weight-at-age, is given in Table 8.2.3. The sampling level in 2006 improved compared to 2005 , considering number measured and aged. In Denmark the provisions in the EU regulation 1639/2001 and the amendment 1581/2004 have been implemented. This provision requires 1 sample per 2000 tonnes landed. This sampling level is lower than the guidelines ( 1 sample per 1000 tonnes) previously used by the HAWG, but as the fishery was carried out in a limited area and a limited season, the recommended sampling level can be regarded as adequate.

In 2006 a total of 601 samples were collected from the Danish vessels taking part in the industrial fisheries in the North Sea in order to estimate species distribution of the industrial landings. The sampling figure for 2005 was 680 samples. The total landings from the Danish small mesh fishery in 2006 were 415000 t (all species) and at the same level as in 2005 (408 000 t ). The recommended sampling levels for species composition were achieved.

### 8.3 Fishery Independent Information

The IBTS (February) sprat indices (no. per hour) in Div. IVb were previously used as an index of abundance of sprat in the North sea. The historical data were revised in 1995 (ICES 1995/Assess:13) and 1999 (ICES 1999/ACFM:12). The IBTS Working Group redefined the sprat index to be calculated as an area weighted mean by rectangles for the entire North Sea sprat stock. New calculations were carried out in 2001 (ICES 2000/D:07). The fishing gear used in the IBTS-survey was standardised in 1983 and the data series from 1984 onwards are considered as comparable (Table 8.3.1).

The IBTS data by rectangle are given in Figure 8.3.1a-c for age groups 1, 2 and 3+. Sprat at all ages were found in the south-east, with the highest concentrations in the more central area of Division IVb and Division IVc. The mean lengths (mm) of age group 1 by rectangles was in general in the range of 62 to 96 mm (Figure 8.3.2). The largest mean length, 102 mm , were in the rectangles 39F0 and 38F0, two rectangles with small landings of sprat

The acoustic surveys for the North Sea Herring in June-July have estimated sprat abundance since 1996 (ICES 2006/LRC:04). No sprat were recorded in the northern part of the North Sea (Figure 8.3.3). The highest abundance and biomass were observed in the south - eastern North Sea. Due to inappropriate coverage of this area during the first period of survey time series, the acoustic estimates are not thought to be representative for the years prior to 2003. In 2004 0 -group sprat ( $<5-6 \mathrm{~cm}$ ) were for the first time recorded by this survey contributing to $34 \%$ of the total abundance It is, however, not clear whether the component of 0 -ringer were recruiting from autumn-spawning sprat or from an early spring-spawning component (ICES 2004/AFM:18). The length distribution indicates that only the largest of this age group have been sampled and the abundance of 0 -group sprat is thus considered an underestimate. In this period no sprat have been reported in the northern areas.

From 2003 the estimates are considered comparable with regards to area covered and are given in the text table below. There was a decrease in numbers and biomass from 2005 to 2006 with the 2004-year class still being the strongest in this period.

|  | Numbers |  |  |  |  | Biomass |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | ${ }^{3+}$ | Sum | 0 | 1 | 2 | ${ }^{3+}$ | Sum |
| 2003 | $\begin{array}{r} 0 \\ 17400 \\ 0 \\ 0 \end{array}$ | $\begin{aligned} & 25292 \\ & 28940 \\ & 70175 \\ & 22889 \end{aligned}$ | $\begin{array}{r} 3984 \\ 5180 \\ 5533 \\ 20480 \end{array}$ | $\begin{array}{r} \hline 339 \\ 99 \\ 1106 \\ 809 \end{array}$ | $\begin{aligned} & \hline 29616 \\ & 51620 \\ & 76814 \\ & 44179 \end{aligned}$ | $\begin{array}{r} \hline 0,0 \\ 19,4 \\ 0,0 \\ 0,0 \end{array}$ | $\begin{aligned} & \hline 198,8 \\ & 266,6 \\ & 479,6 \\ & 166,4 \end{aligned}$ | $\begin{array}{r} \hline 61,3 \\ 71,5 \\ 67,4 \\ 273,0 \end{array}$ | 6,0 | 266,1 |
| 2004 |  |  |  |  |  |  |  |  | 2,1 | 359, |
| 2005 |  |  |  |  |  |  |  |  | 16,8 | 563, |
| 2006 |  |  |  |  |  |  |  |  | 12,7 | 452,1 |

### 8.4 Mean weights-at-age and maturity-at-age

Mean weights (g) at age in the landings in 2006 are presented by quarter in Table 8.4.1. The table includes mean weights-at-age for 1996-2005 for comparison.

Data on maturity by age, mean weight and length by age during the 2006 summer acoustic survey are presented by the PGHERS for the North Sea (ICES 2007/LRC:01) and given in Table 8.4.2. The overall mean weight-at-age for the 1 -group was lower for both immature and mature in 2006 compared to 2005. A higher proportion of the 1-group was mature in 2006 than in the year before, $94 \%$ vs. $57 \%$ by number, respectively. By biomass the proportions were $96 \%$ and $93 \%$.

### 8.5 Recruitment

The 2005 - index of 1-group (2004 year class) was the highest for the whole time-series (see Table 8.3.1), both in absolute and relative terms. The high level of the 1-group in 2005 was seen in most samples and not only confined to few single hauls. In 2007 the IBTS-index for the 2004 year-class (3-group) is still abundant and represents one of the highest 3-group indices for the period. The 1-group index from February 2006 was one of the lowest for the period and the lowest since 1996. The total index was higher than last year, but still lower than the average (1068) for 1984-2007. The 2005 year class appeared as one of the smallest year classes as 1 -ringer in the 2006-IBTS-data. In 2007, the same year class is estimated as the highest 2-group index in the time series.

### 8.6 Data Exploration and Assessment

Assessing the sprat in the North Sea has always been problematic for a variety of reasons. In this section, an elementary exploration of the data is presented to outline the extent of some of these problems. Section 8.7 describes exploratory runs with CSA which has been the assessment procedure employed in later years.

Sprat is a short lived species where the landings are dominated by ages $1-2$ and, to some extent, age 3. Accordingly, the number of observations from each cohort throughout its lifespan is low, making the final estimate of the abundance and mortality of a year class very vulnerable to noise in the data. Furthermore, the fishing mortality is probably not large compared to the assumed natural mortality, so the liberty to model the dynamics of the stock is, to a large extent, limited by the assumption on natural mortality.

The information available at present is landings (in numbers at age) and the IBTS $1^{\text {st }}$ quarter indices (at age). There may also be the potential for the application of acoustic survey data in the future; however, at present, the time-series of this information is of not of sufficient length to justify its use.

## Survey data

IBTS survey indices are available for North Sea sprat for recruits (age 1) and older (age 2+) from 1984 - 2007 (Figure 8.6.1 and Table 8.3.1). The indices for the recruits are highly variable, with a large peak at remarkably regular intervals of 5 years. There is no obvious biological reason for this periodicity. The age $2+$ index also shows periodic fluctuations, but
not as clearly as for the recruits. The correlation between the age 1 IBTS index and the age $2+$ index in the subsequent year (Figure 8.6.2) is weak ( $r^{2}=0.05$ ), and in particular, many large indices at age $2+$ are associated with small indices at age 1 . An analysis using only the age 2 year class, instead of the combined age $2+$ group, gave comparable results.

The survey indices are driven by a few large catches from a small number of stations (Figure 8.6.3), which may explain some of the noise in the survey data. In the extreme year 1989, more than half the total age 1 index was due to a single haul. In more normal years, $40-70 \%$ of the total catch came from the 10 largest hauls (Figure 8.6.3).

Another source of variance in the survey indices may be problems with age reading. Possible variations in the time when the sprat recruit to the survey can add further noise.

The log of the ratio between age 1 and age 2 indices within the same year classes can be used as a metric of the mortality signal in the data, modified by the catchabilities (Figure 8.6.4). The ratio is highly variable and lacks any clear trend, suggesting a high level of noise in the survey data.

## Landing data

The annual landings in numbers at ages 1 and 2, and the corresponding log landing ratio (LLR), are shown in Figure 8.6.5 for the years 1996-2006. The ratios show a marked increase over time due to the increasing trend in landings at age 1 and more stable landings at age 2. These trends may be due to either an increase in total mortality or an increased selection at age 1 in the fishery or by predators. The 2004 year class, which was consistently strong in the survey (Figure 8.6.1), gave rise to relatively large landings first at age 1 and then at age 2, suggesting that the fishery may have targeted this year class. However, the mechanisms behind such behaviour in an industrial fishery are unclear.

## Conclusions with regard to the data

The inferences that can be drawn from these analyses are:

1. The survey data are generally noisy, with an apparent periodic cycle in both age 1 and age $2+$ groups occurring without a clear explanation. However, the correlation between age-group indices in successive years is poor, also likely due to uncertainty in the survey. One major contribution to this noise may be that in most years the indices are driven by a small number of hauls.
2. The log age-index ratio for the IBTS survey is dominated by noise. Hence, there is no clear mortality signal in the survey data.
3. The log landings ratio shows a consistent increasing trend for ages 1-2 since 1996 which is caused by increasing landings at age 1 and constant or slightly decreasing catches at older ages. This could indicate either an increasing F or a change in exploitation pattern with higher preference for age 1. Without some change in selection, the increase in F would have been quite marked.

Further exploration is required including incorporation of acoustic surveys and/or lengthbased assessment methods.

### 8.7 Assessment with CSA

As in previous years, an attempt was made to assess North Sea sprat using the CSA method. The method is described in Section 1.6.2, and in detail in the 2006 HAWG report (ICES CM 2006/ACFM:20). Briefly, a new year-class is entered each year and reduced according to the reported landings (without error) and the assumed natural mortality. Only two ages are
considered in the method; recruits (age 1) and older individuals (ages $2+$ ). The model is fitted to survey indices for the two age groups. The assessment was carried out using the CSAo V3 software obtained through the ICES website (http://www.ices.dk/committe/acfm/wg/asoft/CSA/) with catch and IBTS q1 survey indices from 1996 onwards. An average natural mortality for sprat of 0.75 was assumed, based on the estimated mortalities from an MSVPA analysis for the years 1996-2003 (ICES CM 2005/D:06) raised to the IBTS q1 age-group indices for the appropriate year.

Experience has shown that the survey catchability ratio, s, (i.e. the catchability at the youngest age relative to the older ages, $\mathrm{q}_{1} / \mathrm{q}_{2^{+}}$) required by the model cannot be firmly estimated. The method employed in both this year and previous years has been to scan a range of $s$ values to find the best fit (Figure 8.7.1; note the logarithmic scale for biomass). In the sprat case, the best model fit is obtained at a catchability ratio of 1.075 and this value has been used in the results presented below. However, the terminal year biomass scales extremely strongly with the choice of this parameter (Figure 8.7.1); the dependence of biomass on $s$ is even stronger than exponential in form.

This phenomenon can be understood in terms of the structure of the CSA model. The natural mortality of sprat appears to be significantly higher than the fishing mortality, and thus the relative impact of the catch data on the stock dynamics is low. Thus, while the catches are the only absolute measure of abundance available to the model, their contribution to the dynamics is small. Furthermore, the model interprets these catches literally (i.e. without error) and fits the stock parameters to the indices. In scenarios where the survey-quality is high, the method can be expected to perform well; however, when the survey data are poor and/or conflicting, the model will attempt to eliminate the effect of the catch data by making the biomass extremely large (essentially ignoring this data) and fitting the indices as closely as possible. This appears to be the case with this stock; by changing the catchability parameter, $s$, we are changing the effective weighting of the indices with respect to each other, and thus forcing the model to go to new extremes to fit them correctly.

Thus, given the inability to fix the $s$ value independently of the model and noting the breadth of the minima in the $s$ vs SSQ relationship, the values of the biomass given by the CSA model cannot be considered to have any meaning in absolute terms; rather, they must be considered a measure of the relative abundance of sprat in the North Sea. This philosophy has been employed throughout the remainder of the section.

The input and main results from the key run are presented in Table 8.7.1 and 8.7.2. The CSA model generated a time series of the estimated recruit and mature age-groups and IBTS q1 indices; comparison with the actual indices shows good agreement between the raw data and the fit (Figure 8.7.2). From the size of each individual age-group it is then possible to calculate a time series of the total stock biomass (Figure 8.7.3). The total stock has decreased by approximately $50 \%$ from the ten-year high observed in 2005, and is now at an average value for this period. Exploratory runs performed with different values of $s$ gave qualitatively similar results, both in terms of the quality of fit to the IBTS indices and the dynamics of the total stock biomass, but rescaled.

The fishing mortality estimates from the CSA model are remarkably low without any clear trend (Figure 8.7.4). The trends seen in the log landings ratios (Figure 8.6.5) are not reproduced by CSA. The average fishing mortality over the period 1996-2003 is 0.16 , well below that derived via MSVPA (ICES CM 2005/D:06) which was 0.71 on average over the same period. However, the total stock-biomass, and thus the fishing mortality, $F^{*}$, scale with the catchability ratio, $s$. The discrepancy between the CSA and MSVPA F values can be explained in terms of this scaling.

A retrospective analysis of the stock using the CSA model was limited by the relatively short time series of reliable catch data and only three retrospective runs were feasible (Figure 8.7.5). A strong bias can be seen towards the data point in the terminal year, which had the effect of dragging the total stock biomass upwards. However, the retrospective analyses for the previous two years are in close agreement. This phenomenon is most likely due to unrepresentative values of the 2007 IBTS indices; the increasing trend in these indices in recent years (Figure 8.6.1) tends to contradict the decreasing trend in the biomass as estimated by CSA (Figure 8.7.3).

In conclusion, although the CSA method appears to be a sensible approach to assessing this short-lived species, the noise in the survey data, together with the sensitivity of the model to this noise, make the determination of absolute stock estimates infeasible. Hence, other model formulations should be considered for the future, including the use of more catch-independent data, such as acoustic surveys, to get firmer estimates of the stock. The CSA model does, however, provide a relative estimate of the total stock biomass, and thus is a useful measure of the stock dynamics. The 2007 total-stock-biomass shows a reduction of $50 \%$ from the 10 -year high in 2005.

The HAWG briefly considered preliminary explorations with a length based method (lcs Skagen WD). Again, it is limited what can be inferred from the data, and the results are sensitive to the way the model is conditioned. Besides avoiding problems with ageing, the underlying operating model in lcs is more complex than in CSA, implying that more assumptions are required to estimate the remaining parameters, but also giving more freedom to condition the model according to the insight in the fishery and the stock. Although this approach may be promising, the model is still under development and it would be premature to consider this as an alternative at present. However, the HAWG recommends that ICES be prepared to present IBTS survey indices also by length classes in the future in order to facilitate the use of length-based models of this type.

### 8.8 North Sea Sprat Forecasts

A catch prediction for the assessment year was provided in the past on the basis of a linear regression of catch (as estimated by landings) versus the IBTS sprat index summed over all age groups. The results for 2007 (Figure 8.8.1) indicate a catch in the coming year of 196 kt (agreed TAC for 2007 is 175 kt ). Although such a method has been common in previous years the approach is less than ideal. The relationship between estimated biomass and catch is subject to many different factors such as uncertainty in the stock estimate, recruitment variability and the fact that the TAC is not always taken for this fishery, and is thus not an ideal management strategy.

A framework was developed in an attempt to provide a sounder basis for making short-term forecasts for the North Sea sprat. The CSA model described above was used as the basis for projecting the current estimate of total-stock-biomass forward in time to 2008; although this method has some shortcomings, it can still provide a useful tool in such a scenario. The deterministic equation underpinning the method (Section 1.6.2) was used as described above, but statistical uncertainty was also incorporated by propagating random variations in each of the input variables through the model. For an assumed 2007 catch, and randomly drawn input parameters, it was possible to obtain an estimate of the 2008 total-stock-biomass; by then repeating this process many thousands of times, it was possible to obtain an estimate of both the expected biomass and the confidence intervals surrounding it.

Careful consideration was paid to the variances in the input values. The 2007 North Sea sprat biomass, survey catchability, $\mathrm{q}_{\mathrm{n}}$, and their variances were obtained from the output of the CSA model for an $s$ value of 1.075 and a natural mortality of 0.75 , based on a non-parametric bootstrap performed by the software. An estimate of the variance in the 2007 IBTS sprat age 1
(recruits) index was provided by ICES, based on a similar bootstrapping method. The natural mortality of sprat was assumed to be log-normally distributed with a variance based on the analysis of values provided by the SGMSNS 2005 MSVPA (ICES CM 2005/D:06). A typical exploitation pattern for the fishery was estimated based on the output of the CSA model from 1996-2006 and was used to determine the mean weight-in-catch. Finally, the sprat recruitment for 2008 was assumed to log normally distributed about the geometric mean, also obtained from the output of the CSA model.

The 2008 sprat forecast is a linearly decreasing function of total catch taken during 2007 (Figure 8.8.2), as expected from the CSA model; the slope of the line agrees with that predicted analytically from CSA. The variance about the median value is log-normally distributed as expected and analysis shows it to be driven by the uncertainty in the coming year's recruitment in the first instance, and that in the 2007 stock biomass in the second.

The interpretation of these results should be tempered by our understanding of the limitations of the CSA model. The CSA method produces a relative estimate of the total stock biomass whose scaling is driven by the $s$ parameter, which was assumed to be constant in the forecast model above. The choice of $s$ is again critical as it will scale the 2007 biomass and thus change the impact of removing a fixed tonnage from the stock. Exploratory runs were performed for different $s$ values by re-estimating all parameters from scratch, based on the appropriate CSA model run. Forecasts of the 2008 biomass based on $s=0.7$ show an increased dependence on catch (Figure 8.8.3), while those made with $s=1.3$ were almost independent of catch.

The utility of this forecast method is thus limited by the validity of the CSA method. However, further refinements of the core model, as discussed above, can potentially improve the quality of the North Sea sprat assessment, and thus also of this forecast approach.

### 8.9 Quality of the Assessment

Due to the nature of the methods employed, the assessment of this stock is heavily dependent on the quality of the IBTS sprat indices. An investigation of the structure of these metrics revealed significant questions about their reliability; it was found that the ten largest hauls commonly comprised $40-70 \%$ of the index, and in some exceptional years more than $50 \%$ of the index was driven by a single haul. In addition, HAWG is aware of problems in the IBTS with the timing of recruitment to the survey; some sprat that hatch in autumn may not be fully recruited by February in the next year.

The quality of the assessment is also severely limited by the inability of the CSA model to provide an absolute estimate of the stock biomass. The key to this method is the catchability ratio, $s$; whilst it is possible to obtain a "best-guess" value via SSQ profiling, the minimum is extremely broad and the resulting estimate is thus uncertain. HAWG is not aware of any method to estimate this factor independently of the model and must thus rely on the fitted value; the parameter thus reflects uncertainties in the survey indices, rather than the fundamental physical and/or ecological processes it is supposed to capture. The biomass in the terminal year is shown to be an extremely strong function of this parameter, and thus the assessment of the absolute 2007 biomass is meaningless.

Investigations have shown that the relative trends in the assessment time-series produced by CSA are independent of the value of the $s$ parameter. The fit of the model to IBTS sprat indices is generally good, and in this regard, CSA is able to provide useful information about the state of the stock in relative terms.

### 8.10 Management Considerations

The size of the North Sea sprat stock is mostly driven by the recruiting year class. Thus, the fishery in a given year will be dependent on that year's incoming year-class and only in-year catch forecasts are available. The sprat stock in the previous two years has been dominated by the very strong 2004 year class; this group has now passed through the fishery and the estimated biomass has returned to an average value. The 2006 year class, as indicated by the 2007 IBTS q1 age 1 index, is marginally below the ten-year geometric mean. The IBTS index describing the recruitment of the 2005 year class (i.e. the 2006 age 1 index) was particularly weak, but the effect does not appear to have propagated through into the 2007 age 2+ index, which shows an increase to one of its highest observed levels. The combined IBTS index for 2007 is high ( $80^{\text {th }}$ percentile) but contradicts the relative estimate of biomass obtained from the CSA model ( $50^{\text {th }}$ percentile). Due to inconsistencies in the age structure of the 2007 IBTS indices (particularly the absence of any effect due to the weak recruitment of the 2005 yearclass), HAWG believes that the relative estimate obtained from the model is the more reliable of the two metrics.

There are indications that larvae from autumn spawning will over-winter as larvae and metamorphose the year after. A better understanding of the stock structure and spatial distribution of sprat in the North Sea, the spawning seasons and recruitment from a possible autumn spawning is required.

Uncertainties in both the assessment method and the survey indices make the current understanding of this stock extremely poor. HAWG recommends that the detailed study of improved or alternative assessment methods (e.g. length-based assessment) and the use of additional information sources (e.g. acoustic surveys, catch per unit effort) are required in order improve our level of understanding and ability to adequately manage this stock.

Table 8.1.1. Sprat in the North Sea. Catches (' 000 t t 1996-2006. See ICES CM 2006/ACFM:20 for earlier catch data. Catch in fords of western Norway excluded. (Data provided by Working Group members except where indicated). These figures do not in all cases correspond to the official statistics and cannot be used for management purposes.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Division IVa West (North Sea) stock |  |  |  |  |  |  |  |  |  |  |  |
| Denmark |  |  |  | 0.7 |  | 0.1 | 1.1 |  | 0.0 |  | 0.0 |
| Netherlands |  |  |  |  |  |  |  |  |  |  |  |
| Norway |  |  |  |  |  |  |  |  |  |  |  |
| Sweden |  |  |  |  |  | 0.1 |  |  |  |  |  |
| UK(Scotland) |  |  |  |  |  |  |  |  |  |  |  |
| Total | 0.0 | 0.0 | 0.0 | 0.7 | 0.0 | 0.2 | 1.1 |  | 0.0 | 0.0 | 0.0 |
| Division IVa East (North Sea) stock |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 0.3 |  |  |  |  |  |  |  |  |  | 0.0 |
| Norway |  |  |  |  |  |  |  |  |  |  |  |
| Sweden |  |  |  |  |  |  |  |  |  |  |  |
| Total | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  |  |  | 0.0 |
| Division IVb West |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 1.8 | 82.2 | 21.1 | 13.2 | 18.8 | 11.1 | 16.3 | 22.0 | 53.8 | 53.3 | 8.0 |
| Norway | 1.9 | 2.3 |  |  |  | 0.9 | 0.0 |  |  |  |  |
| UK(Engl.\&Wales) |  |  |  |  |  |  |  |  |  |  |  |
| UK(Scotland) |  |  |  | 0.8 |  |  |  |  |  |  |  |
| Total | 3.7 | 84.5 | 21.1 | 14.0 | 18.8 | 12.0 | 16.3 | 22.0 | 53.8 | 53.3 | 8.0 |
| Division IVb East |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 74.7 | 10.9 | 98.2 | 147.1 | 144.1 | 132.9 | 109.8 | 130.9 | 122.2 | 150.7 | 71.5 |
| Germany |  |  |  |  |  |  |  |  |  |  |  |
| Norway | 50.9 | 0.8 | 15.3 | 13.1 | 0.9 | 5.0 |  |  | 0.1 |  | 0.8 |
| Sweden | 0.5 |  | 1.7 | 2.1 |  | 1.4 |  |  |  | 0.0 |  |
| UK(Scotland) |  |  |  | 0.6 | 0.0 |  |  |  |  |  |  |
| Total | 126.1 | 11.7 | 115.2 | 162.9 | 145.0 | 139.3 | 109.8 | 131.0 | 122.2 | 150.7 | 72.3 |
| Division IVc |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 3.9 | 5.7 | 11.8 | 3.3 | 28.2 | 13.1 | 14.8 | 22.3 | 16.8 | 2.0 | 23.8 |
| France |  |  |  |  |  |  | 0.0 |  |  |  |  |
| Netherlands |  |  |  | 0.2 |  |  | 0.0 |  |  |  |  |
| Norway |  | 0.1 | 16.0 | 5.7 | 1.8 | 3.6 | 0.0 |  |  |  | 9.0 |
| UK(Engl.\& | 2.6 | 1.4 | 0.2 | 1.6 | 2.0 | 2.0 | 1.6 | 1.3 | 1.5 | 1.6 | 0.5 |
| Total | 6.5 | 7.2 | 28.0 | 10.8 | 32.0 | 18.7 | 16.4 | 23.6 | 18.3 | 3.6 | 33.4 |
| Total North Sea |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 80.7 | 98.8 | 131.1 | 164.3 | 191.1 | 157.2 | 142.0 | 175.2 | 192.7 | 206.0 | 103.4 |
| France | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  |  |  |  |  |
| Germany | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  |  |  |  |  |
| Netherland: | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 |  |  |  |  |  |
| Norway | 52.8 | 3.2 | 31.3 | 18.8 | 2.7 | 9.5 | 0.0 | 0.0 | 0.1 | 0.0 | 9.8 |
| Sweden | 0.0 | 0.0 | 0.0 | 2.7 | 0.0 | 1.4 |  |  |  |  |  |
| UK(Engl.\& | 2.6 | 1.4 | 0.2 | 1.6 | 2.0 | 2.0 | 1.6 | 1.3 | 1.5 | 1.6 | 0.5 |
| UK(Scotla | 0.0 | 0.0 | 0.0 | 0.8 | 0.0 | 0.0 |  |  |  |  |  |
| Total | 136.1 | 103.4 | 162.6 | 188.4 | 195.9 | 170.1 | 143.6 | 176.5 | 194.3 | 207.7 | 113.7 |

Table 8.1.2. Sprat catches ( ' 000 t ) in the fjords of western Norway, 1985-2006.

| 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 7.1 | 2.2 | 8.3 | 5.3 | 2.4 | 2.7 | 3.2 | 3.8 | 1.9 | 5.3 | 3.7 | 3.3 | 3.1 | 2.5 | 3.3 | 2.6 | 1.4 | 1.1 | 2.2 | 0.4 | 1.2 | 1.3 |

${ }^{1}=$ preliminary

Table 8.1.3. Sprat in the North Sea. Catches (tonnes) by quarter*. Catches in fjords of Western Norway excluded.


Table 8.2.1. North Sea sprat. Species composition in the Danish sprat fishery in tonnes and percentage of the total catch. Data is reported for 1998-2006.

|  | Year | Sprat | Herring | Horse-mackerel | Whiting | Haddock | ackerel | Cod | Sandeel | Other species | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Tonnes | 1998 | 129,315 | 11,817 | 573 | 673 | 6 | 220 | 11 | 2,174 | 1,188 | 145,978 |
| Tonnes | 1999 | 157,003 | 7,256 | 413 | 1,088 | 62 | 321 | 7 | 4,972 | $635{ }^{\text { }}$ | 171,757 |
| Tonnes | 2000 | 188,463 | 11,662 | 3,239 | 2,107 | 66 | 766 | 4 | 423 | 1,911 ${ }^{\text {r }}$ | 208,641 |
| Tonnes | 2001 | 136,443 | 13,953 | 67 | 1,700 | 223 | 312 | 4 | 17,020 | 1,142 ${ }^{\prime \prime}$ | 170,862 |
| Tonnes | 2002 | 140,568 | 16,644 | 2,078 | 2,537 | 27 | 715 | 0 | 4,102 | $800{ }^{*}$ | 167,471 |
| Tonnes | 2003 | 172,456 | 10,244 | 718 | 1,106 | 15 | 799 | 11 | 5,357 | 3,509 | 194,214 |
| Tonnes | 2004 | 179,944 | 10,144 | 474 | 334 |  | 4,351 | 3 | 3,836 | 1,821 | 200,906 |
| Tonnes | 2005 | 201,331 | 21,035 | 2,477 | 545 | 4 | 1,009 | 16 | 6,859 | 974 | 234,250 |
| Tonnes | 2006 | 103,236 | 8,983 | 577 | 343 | 25 | 905 | 4 | 5,384 | $576{ }^{\text {² }}$ | 120,033 |
| Percent | 1998 | 88.6 | 8.1 | 0.4 | 0.5 | 0.0 | 0.2 | 0.0 | 1.5 | 0.8 | 100.0 |
| Percent | 1999 | 91.4 | 4.2 | 0.2 | 0.6 | 0.0 | 0.2 | 0.0 | 2.9 | 0.4 | 100.0 |
| Percent | 2000 | 90.3 | 5.6 | 1.6 | 1.0 | 0.0 | 0.4 | 0.0 | 0.2 | 0.9 | 100.0 |
| Percent | 2001 | 79.9 | 8.2 | 0.0 | 1.0 | 0.1 | 0.2 | 0.0 | 10.0 | 0.7 | 100.0 |
| Percent | 2002 | 83.9 | 9.9 | 1.2 | 1.5 | 0.0 | 0.4 | 0.0 | 2.4 | 0.5 | 100.0 |
| Percent | 2003 | 88.8 | 5.3 | 0.4 | 0.6 | 0.0 | 0.4 | 0.0 | 2.8 | 1.8 | 100.0 |
| Percent | 2004 | 89.6 | 5.0 | 0.2 | 0.2 | 0.0 | 2.2 | 0.0 | 1.9 | 0.9 | 100.0 |
| Percent | 2005 | 85.9 | 9.0 | 1.1 | 0.2 | 0.0 | 0.4 | 0.0 | 2.9 | 0.4 | 100.0 |
| Percent | 2006 | 86.0 | 7.5 | 0.5 | 0.3 | 0.0 | 0.8 | 0.0 | 4.5 | 0.5 | 100.0 |

Table 8.2.2 North Sea Sprat. Catch in numbers (millions) by quarter and by age 1996-2006.

| Year | Quarter | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0 | 1 | 2 | 3 | 4 | 5+ | Total |
| 1996 | 1 |  | 524.7 | 4,615.4 | 2,621.9 | 316.4 | 11.3 | 8,089.7 |
|  | 2 |  | 1.9 | 241.5 | 32.7 | 15.5 | 0.3 | 291.9 |
|  | 3 |  | 400.5 | 100.7 | 22.9 | 0.3 |  | 524.5 |
|  | 4 |  | 1,190.7 | 1,069.0 | 339.6 | 5.6 |  | 2,604.8 |
| Total |  |  | 2,117.9 | 6,026.6 | 3,017.0 | 337.8 | 11.5 | 11,510.8 |
| 1997 | 1 |  | 74.4 | 314.0 | 229.2 | 55.3 | 2.5 | 675.4 |
|  | 2 |  | 11.3 | 47.8 | 34.9 | 8.4 | 0.4 | 102.9 |
|  | 3 |  | 1,991.9 |  |  |  |  | 1,991.9 |
|  | 4 | 127.6 | 3,597.2 | 996.2 | 117.8 | 58.1 | 0.0 | 4,896.9 |
|  | Total | 127.6 | 5,674.8 | 1,358.1 | 381.9 | 121.8 | 2.8 | 7,667.1 |
| 1998 | 1 |  | 683.2 | 537.2 | 18.3 | 0.1 |  | 1,238.8 |
|  | 2 |  | 70.9 | 55.3 | 1.8 |  |  | 127.9 |
|  | 3 | 74.2 | 3,356.6 | 693.3 |  |  |  | 4,124.2 |
|  | 4 | 772.4 | 4,822.4 | 2,295.1 | 483.5 | 39.5 |  | 8,412.8 |
|  | Total | 846.6 | 8,933.1 | 3,580.9 | 503.6 | 39.6 |  | 13,903.7 |
| 1999 | 1 |  | 728.1 | 2,226.0 | 554.2 | 86.6 | 9.2 | 3,604.2 |
|  | 2 |  | 38.6 | 58.4 | 18.1 | 2.6 |  | 117.7 |
|  | 3 |  | 12,919.0 | 38.9 |  |  |  | 12,957.8 |
|  | 4 | 105.0 | 2,143.2 | 211.5 |  |  |  | 2,459.7 |
|  | Total | 105.0 | 15,828.9 | 2,534.8 | 572.3 | 89.2 | 9.2 | 19,139.5 |
| 2000 | 1 |  | 559.2 | 3,177.3 | 797.5 | 247.5 | 72.0 | 4,853.7 |
|  | 2 |  | 6.8 | 107.4 | 60.1 | 12.8 | 0.5 | 187.6 |
|  | 3 |  | 9,928.9 | 1,111.9 | 77.8 |  |  | 11,118.6 |
|  | 4 |  | 1,153.7 | 129.2 | 9.0 |  |  | 1,291.9 |
|  | Total |  | 11,648.7 | 4,525.8 | 944.4 | 260.3 | 72.6 | 17,451.8 |
| 2001 | 1 |  | 746.3 | 3,197.7 | 1,321.9 | 22.2 |  | 5,023.1 |
|  | 2 |  | 15.9 | 66.2 | 26.1 |  |  | 108.2 |
|  | 3 | 0.4 | 3,338.8 | 299.9 |  |  |  | 3,559.1 |
|  | 4 | 1,205.0 | 4,178.7 | 1,224.6 | 261.9 |  |  | 6,651.4 |
|  | Total | 1,205.4 | 8,279.8 | 4,788.4 | 1,609.9 | 22.2 |  | 15,341.7 |
| 2002 | 1 | 0.0 | 104.7 | 400.3 | 30.2 | 11.2 |  | 546.4 |
|  | 2 | 0.0 | 13.7 | 27.9 | 2.4 | 0.6 |  | 44.6 |
|  | 3 | 40.9 | 5,745.6 | 582.1 | 42.3 | 4.1 |  | 6,415.0 |
|  | 4 | 415.0 | 4,578.0 | 626.2 | 119.8 | 3.1 | - | 5,742.1 |
|  | Total | 455.9 | 10,441.9 | 1,636.5 | 194.8 | 19.0 |  | 12,748.1 |
| 2003 | 1 | 0.0 | 1,953.9 | 1,218.9 | 85.3 | 11.3 | 0.0 | 3,269.3 |
|  | 2 | 0.0 | 41.8 | 46.3 | 4.7 | 0.6 | $0.0^{\text {r }}$ | 93.3 |
|  | 3 | 1.1 | 3,481.3 | 772.0 | 42.9 | 0.0 | 0.0 " | 4,297.2 |
|  | 4 | 539.3 | 7,051.8 | 1,115.1 | 93.8 | 36.5 | 21.9 " | 8,858.4 |
|  | Total | 540.4 | 12,528.7 | 3,152.3 | 226.6 | 48.4 | $21.9{ }^{\prime \prime}$ | 16,518.2 |
| 2004 | 1 | 0.0 | 16.5 | 214.0 | 26.3 | 1.6 | 0.6 | 259.0 |
|  | 2 | 0.0 | 22.1 | 14.9 | 3.0 | 0.1 | $0.0^{\prime \prime}$ | 40.1 |
|  | 3 | 210.0 | 3,661.9 | 558.2 | 31.4 | 0.0 | $0.0{ }^{\text {F }}$ | 4,461.5 |
|  | 4 | 15,674.4 | 5,582.8 | 632.1 | 59.2 | 0.0 | $0.0^{*}$ | 21,948.5 |
|  | Total | 15,884.4 | 9,283.2 | 1,419.2 | 119.8 | 1.8 | 0.6 | 26,709.1 |
| 2005 | 1 | 0.0 | 2,476.5 | 268.5 | 13.8 | 2.2 | 0.0 | 2,761.1 |
|  | 2 | 0.0 | 499.6 | 23.4 | 4.3 | 4.9 | $0.0{ }^{\prime \prime}$ | 532.1 |
|  | 3 | 0.0 | 11,920.2 | 192.3 | 7.6 | 0.0 | $0.0^{\prime \prime}$ | 12,120.0 |
|  | 4 | 302.5 | 7,467.9 | 191.1 | 0.0 | 0.0 | $0.0^{\prime \prime}$ | 7,961.6 |
| Total |  | 302.5 | 22,364.3 | 675.3 | 25.7 | 7.0 | 0.0 | 23,374.8 |
| 2006 | 1 | 0.0 | 1,559.2 | 5,119.1 | 95.7 | 2.3 | $0.0{ }^{\prime \prime}$ | 6,776.2 |
|  | 2 | 0.0 | 5.8 | 21.5 | 0.2 | 0.0 | $0.0{ }^{\text {² }}$ | 27.4 |
|  | 3 | 0.0 | 3,077.8 | 625.0 | 129.1 | 0.0 | $0.0{ }^{\text {² }}$ | 3,831.9 |
|  | 4 | 0.0 | 2,048.5 | 416.0 | 85.9 | 0.0 | 0.0 " | 2,550.4 |
|  | Total | 0.0 | 6,691.2 | 6,181.6 | 310.8 | 2.3 | 0.0 | 13,185.9 |

Table 8.2.3. North Sea Sprat. Sampling for biological samples in 2006.

| Country | Quarter | Landings <br> ('000 tonnes) | No. <br> samples | No. <br> measured | No. <br> aged |
| :--- | :---: | ---: | ---: | ---: | ---: |
| Denmark | 1 | 36.63 | 17 | 2399 | 2399 |
|  | 2 | 0.19 | 0 | 0 | 0 |
|  | 3 | 39.95 | 9 | 918 | 918 |
|  | 4 | 26.59 | 1 | 120 | 0 |
|  | Total | 103.36 | 27 | 3437 | 3317 |
| UK(England) | 1 | 0.54 | 0 | 0 | 0 |
|  | 2 |  |  |  |  |
|  | 3 |  |  |  |  |
|  | 4 |  |  |  |  |
| Norway | Total | 0.54 | 0 | 0 | 0 |
|  | 1 | 9.81 | 6 | 600 | 450 |
|  | 2 |  |  |  |  |
| Total North Sea |  |  |  |  |  |

Table 8.3.1 North Sea sprat. Abundance indices by age from IBTS (February) from 1984-2007

| Year | Age |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5+ Total |
| 1984 | 233.758 | 329.003 | 39.608 | 6.200 | $0.292 \times 608.861$ |
| 1985 | 376.098 | 195.479 | 26.757 | 3.803 | $0.354{ }^{\text { }} 602.491$ |
| 1986 | 44.188 | 73.538 | 22.010 | 1.233 | $0.243^{*} 141.212$ |
| 1987 | 542.236 | 66.279 | 19.144 | 1.924 | $0.240^{*} 629.823$ |
| 1988 | 98.606 | 884.065 | 61.800 | 6.991 | $0.000{ }^{*} 1051.462$ |
| 1989 | 2314.218 | 476.292 | 271.849 | 22.010 | $1.647^{*} 3086.016$ |
| 1990 | 234.942 | 451.979 | 102.164 | 28.063 | $2.219^{\prime \prime} 819.367$ |
| 1991 | 676.784 | 93.381 | 23.330 | 2.631 | $0.118 * 796.244$ |
| 1992 | 1060.780 | 297.691 | 43.248 | 7.234 | $0.531{ }^{*} 1409.484$ |
| 1993 | 1066.829 | 568.530 | 118.416 | 6.074 | $0.338{ }^{\text {² }} 1760.187$ |
| 1994 | 2428.357 | 938.159 | 92.161 | 3.593 | $0.504{ }^{\text { }} 3462.774$ |
| 1995 | 1224.891 | 1036.404 | 87.329 | 2.516 | $0.764{ }^{\text {² }} 2351.904$ |
| 1996 | 186.131 | 383.534 | 146.839 | 18.284 | $0.744^{\prime \prime} 735.532$ |
| 1997 | 591.862 | 411.953 | 179.551 | 15.522 | $2.239{ }^{*} 1201.127$ |
| 1998 | 1171.050 | 1456.508 | 305.908 | 15.753 | $3.381{ }^{\text {² }} 2952.600$ |
| 1999 | 2534.528 | 562.098 | 80.347 | 4.828 | $0.445^{\text {² }} 3182.246$ |
| 2000 | 1058.204 | 851.581 | 274.711 | 43.887 | $0.882^{*} 2229.265$ |
| 2001 | 883.058 | 1057.001 | 185.466 | 17.548 | $0.345^{*} 2143.418$ |
| 2002 | 1152.328 | 812.450 | 91.631 | 11.931 | $0.375^{\text {² }} 2068.715$ |
| 2003 | 1842.261 | 309.918 | 44.491 | 1.022 | $0.040^{*} 2197.732$ |
| 2004 | 1593.892 | 495.702 | 78.243 | 3.498 | $1.536{ }^{*} 2172.871$ |
| 2005 | 3053.458 | 267.892 | 36.385 | 0.868 | 0.000 * 3358.603 |
| 2006 | 421.803 | 1212.870 | 92.378 | 8.262 | $0.072{ }^{\text { }} 1735.385$ |
| 2007 | 934.532 | 1772.646 | 293.834 | 13.572 | $0.026^{*} 3014.610$ |

Table 8.4.1 North Sea Sprat. Mean weight (g) by quarter and by age for 1996-2006.

| Year | Quarter | Age |  |  |  |  |  | SOP <br> Tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0 | 1 | 2 | 3 | 4 | 5+ |  |
| 1996 | 1 |  | 3.9 | 9.3 | 14.9 | 15.3 | 16.1 | 88,807 |
|  | 2 |  | 6.9 | 8.4 | 11.6 | 20.0 | 15.2 | 2,735 |
|  | 3 |  | 11.6 | 14.2 | 18.2 | 21.5 |  | 6,501 |
|  | 4 |  | 12.1 | 15.9 | 17.2 | 20.5 |  | 37,359 |
| Weighted mean |  |  | 10.0 | 10.5 | 15.1 | 15.6 | 16.0 | 135,401 |
| 1997 | 1 |  | 8.0 | 10.0 | 15.0 | 17.0 | 19.0 | 8,161 |
|  | 2 |  | 8.0 | 10.0 | 15.0 | 17.0 | 19.0 | 1,243 |
|  | 3 |  | 14.2 |  |  |  |  | 28,285 |
|  | 4 | 3.7 | 11.9 | 16.4 | 19.1 | 19.6 |  | 63,083 |
| Weighted mean |  | 3.7 | 12.7 | 14.7 | 16.3 | 18.2 | 19.0 | 100,772 |
| 1998 | 1 |  | 5.6 | 6.0 | 8.7 | 15.0 |  | 7,232 |
|  | 2 |  | 5.6 | 6.0 | 8.3 |  |  | 743 |
|  | 3 | 3.7 | 14.7 | 15.3 |  |  |  | 60,149 |
|  | 4 | 4.1 | 10.6 | 13.8 | 16.3 | 14.6 |  | 94,173 |
| Weighted mean |  | 4.0 | 11.7 | 12.8 | 16.0 | 14.7 |  | 162,297 |
| 1999 | 1 |  | 3.3 | 8.7 | 12.5 | 14.4 | 16.3 | 30,168 |
|  | 2 |  | 3.1 | 10.1 | 13.6 | 15.4 |  | 993 |
|  | 3 |  | 10.0 | 18.3 |  |  |  | 129,383 |
|  | 4 | 4.4 | 11.0 | 14.4 |  |  |  | 27,126 |
| Weighted mean |  | 4.4 | 9.8 | 9.4 | 12.5 | 14.4 | 16.3 | 187,670 |
| 2000 | 1 |  | 4.2 | 10.1 | 10.7 | 10.2 | 10.5 | 46,192 |
|  | 2 |  | 3.3 | 9.0 | 10.2 | 12.8 | 10.5 | 1,767 |
|  | 3 |  | 11.9 | 11.9 | 11.0 |  |  | 132,563 |
|  | 4 |  | 11.9 | 11.9 | 11.0 |  |  | 15,403 |
| Weighted mean |  |  | 11.6 | 10.6 | 10.7 | 10.3 | 10.5 | 195,925 |
| 2001 | 1 |  | 3.3 | 9.7 | 12.9 | 16.5 |  | 50,794 |
|  | 2 |  | 3.3 | 10.3 | 12.9 |  |  | 1,071 |
|  | 3 | 4.0 | 12.0 | 15.3 |  |  |  | 44,656 |
|  | 4 | 3.8 | 11.6 | 12.6 | 19.1 |  |  | 73,444 |
| Weighted mean |  | 3.8 | 11.0 | 10.8 | 13.9 | 16.5 |  | 169,967 |
| 2002 | 1 |  | 7.0 | 12.0 | 14.0 | 13.0 |  | 61,057 |
|  | 2 |  | 5.3 | 11.2 | 12.5 | 12.4 |  | 4,231 |
|  | 3 | 2.0 | 10.9 | 15.0 | 15.0 | 24.0 |  | 721,732 |
|  | 4 | 3.9 | 12.0 | 15.0 | 15.7 | 24.0 |  | 679,018 |
| Weighted mean |  | 3.7 | 11.2 | 13.4 | 14.9 | 14.8 |  | 1,466,038 |
| 2003 | 1 |  | 3.6 | 9.4 | 11.0 | 15.0 |  | 19,599 |
|  | 2 |  | 3.1 | 9.9 | 11.0 | 15.0 |  | 648 |
|  | 3 | 3.0 | 13.0 | 16.0 | 13.0 |  |  | 58,169 |
|  | 4 | 4.6 | 10.8 | 14.8 | 16.9 | 15.0 | 18.0 | 97,670 |
| Weighted mean |  | 4.6 | 10.3 | 12.9 | 13.8 | 15.0 | 18.0 | 176,085 |
| 2004 | 1 |  | 3.6 | 10.3 | 13.8 | 16.6 | 16.1 | 2,663 |
|  | 2 |  | 6.0 | 8.5 | 7.3 | 10.2 |  | 282 |
|  | 3 | 4.5 | 11.9 | 17.0 | 20.0 |  |  | 54,639 |
|  | 4 | 4.0 | 11.4 | 14.6 | 18.3 |  |  | 136,653 |
| Weighted mean |  | 4.0 | 11.0 | 10.9 | 14.5 | 16.8 | $16.1{ }^{\prime}$ | 194,238 |
| 2005 | 1 |  | 4.6 | 8.9 | 12.1 | 16.0 |  | 13,995 |
|  | 2 |  | 4.8 | 6.5 | 9.8 | 10.0 |  | 2,641 |
|  | 3 |  | 8.9 | 9.9 | 18.6 |  |  | 107,531 |
|  | 4 | 4.1 | 10.7 | 12.0 |  |  |  | 83,515 |
| Weighted mean |  | 4.1 | 8.9 | 10.0 | 13.6 | 11.8 |  | 207,682 |
| 2006 | 1 |  | 4.3 | 7.7 | 9.6 | 13.0 |  | 47,293 |
|  | 2 |  | 3.7 | 8.1 | 11.2 |  |  | 198 |
|  | 3 |  | 9.8 | 12.5 | 16.1 |  |  | 40,053 |
|  | 4 |  | 9.8 | 12.5 | 16.1 |  |  | 26,658 |
| Weighted mean |  |  | 8.5 | 8.5 | 14.1 | 13.0 |  | 114,202 |

Table 8.4.2. North Sea sprat. Abundance, biomass, mean weight and length by age, maturity for the area east and west of $3^{\circ} \mathrm{E}$ and for the total North Sea.


Table 8.7.1. North Sea sprat. Input data to the CSA model. Catch in numbers (CatRec and CatFull), IBTS q1 abundance indices for age 1 (Irec) and age $2+$ (Ifull), mean weights in the stock of recruits (Wrec) and mature individuals (Wfull), the catchability ratio (Srat) and the natural mortality, (M). Catches for the 2007 year are set to zero, as they are not used by the model. The 2007 lfull index is used as a fitting parameter in this method but the Irec index for 2007 is not.

| Year | CatRec | CatFull | Irec | Ifull | Wrec | Wfull | Srat | M |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1996 | 2118 | 9393 | 186 | 549 | 4.5 | 9.67 | 1.075 | 0.75 |
| 1997 | 5675 | 1865 | 592 | 609 | 4.5 | 9.67 | 1.075 | 0.75 |
| 1998 | 8933 | 4124 | 1171 | 1782 | 4.5 | 9.67 | 1.075 | 0.75 |
| 1999 | 15829 | 3206 | 2535 | 648 | 4.5 | 9.67 | 1.075 | 0.75 |
| 2000 | 11649 | 5803 | 1058 | 1171 | 4.5 | 9.67 | 1.075 | 0.75 |
| 2001 | 8280 | 6420 | 883 | 1260 | 4.5 | 9.67 | 1.075 | 0.75 |
| 2002 | 10442 | 1850 | 1152 | 916 | 4.5 | 9.67 | 1.075 | 0.75 |
| 2003 | 12529 | 3449 | 1842 | 357 | 4.5 | 9.67 | 1.075 | 0.75 |
| 2004 | 9283 | 1542 | 1594 | 579 | 4.5 | 9.67 | 1.075 | 0.75 |
| 2005 | 22364 | 708 | 3053 | 305 | 4.5 | 9.67 | 1.075 | 0.75 |
| 2006 | 6691 | 6495 | 422 | 1314 | 4.5 | 9.67 | 1.075 | 0.75 |
| 2007 | 0 | 0 | 847 | 1841 | 4.5 | 9.67 | 1.075 | 0.75 |

Table 8.7.2 North Sea sprat. Summarised output from the CSA model showing the number of recruited (RecN) and mature (FullN) individuals, the total stock biomass (TSBiom), the estimated fishing mortality ( $\mathrm{F}^{*}$ ), the harvest rates for recruits (HRrec) and mature individuals (HRfull), the catches of recruits (CatRec) and the fully recruited (CatFull), the catchability ratio (Sratio) and natural mortality (M) The table also gives the mature individual catchability (q), the sum of squares of error (SSQ), and the root-mean-square error (RMS).

| Year | RecN | Fulln | TSBiom | F* | HRrec | HRfull | CatRec | CatFull | Sratio | M |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 14821.9 | 75801.6 | 799700 | 0.204 | 0.143 | 0.124 | 2118 | 9393 | 1.075 | 0.75 |
| 1997 | 86757.5 | 34896.1 | 727854 | 0.095 | 0.065 | 0.053 | 5675 | 1865 | 1.075 | 0.75 |
| 1998 | 79277.3 | 52283 | 862324 | 0.156 | 0.113 | 0.079 | 8933 | 4124 | 1.075 | 0.75 |
| 1999 | 183619 | 53170.7 | 1340447 | 0.124 | 0.086 | 0.06 | 15829 | 3206 | 1.075 | 0.75 |
| 2000 | 83347.8 | 98769 | 1330161 | 0.15 | 0.14 | 0.059 | 11649 | 5803 | 1.075 | 0.75 |
| 2001 | 59471.4 | 74031.3 | 983504 | 0.175 | 0.139 | 0.087 | 8280 | 6420 | 1.075 | 0.75 |
| 2002 | 55232.6 | 52959.1 | 760661 | 0.181 | 0.189 | 0.035 | 10442 | 1850 | 1.075 | 0.75 |
| 2003 | 87975.3 | 42657.9 | 808391 | 0.196 | 0.142 | 0.081 | 12529 | 3449 | 1.075 | 0.75 |
| 2004 | 70197.9 | 50725.3 | 806404 | 0.14 | 0.132 | 0.03 | 9283 | 1542 | 1.075 | 0.75 |
| 2005 | 277968.1 | 49680.1 | 1731263 | 0.108 | 0.08 | 0.014 | 22364 | 708 | 1.075 | 0.75 |
| 2006 | 34701.1 | 138912.9 | 1499443 | 0.117 | 0.193 | 0.047 | 6691 | 6495 | 1.075 | 0.75 |
| 2007 | 60110.5 | 72946.9 | 975893 | 0 | 0 | 0 | 0 | 0 | 1.075 | 0.75 |
| $q=$ | 1.31E-02 |  |  |  |  |  |  |  |  |  |
| SSQ = | 4.02E+00 |  |  |  |  |  |  |  |  |  |
| RMS = | 3.65E-01 |  |  |  |  |  |  |  |  |  |

## Sprat catches 2006, 1st Quarter



Figure 8.1.1a. Sprat catches (in tonnes) in the North Sea and Div. IIIa in 2006 by statistical rectangle. Working group estimates. First quarter.

Sprat catches 2006, 2nd Quarter


Figure 8.1.1b. Sprat catches (in tonnes) in the North Sea and Div. IIIa in 2006 by statistical rectangle. Working group estimates. Second quarter.

## Sprat catches 2006, 3rd Quarter



Figure 8.1.1c. Sprat catches (in tonnes) in the North Sea and Div. IIIa in 2006 by statistical rectangle. Working group estimates. Third quarter.

## Sprat catches 2006, 4th Quarter



Figure 8.1.1d. Sprat catches (in tonnes) in the North Sea and Div. IIIa in 2006 by statistical rectangle. Working group estimates. Fourth quarter.

Sprat catches 2006, All Quarters


Figure 8.1.2. Sprat catches (in tonnes) in the North Sea and Div. IIIa in 2006 by statistical rectangles. Working group estimates.

## Sprat 1-ringers IBTS 1st Quarter 2007



Figure 8.3.1a. Sprat. Distribution of age group 1 in the IBTS (February) 2007 in the North Sea and Division IIIa (Mean number per hour per rectangle).

## Sprat 2-ringers IBTS 1st Quarter 2007



Figure 8.3.1b. Sprat. Distribution of age group 2 in the IBTS (February) 2007 in the North Sea and Division IIIa (Mean number per hour per rectangle).

Sprat 3+ ringers IBTS 1st Quarter 2007


Figure 8.3.1c. Distribution of age group 3+ in the IBTS (February) 2007 in the North Sea and Division IIIa (Mean number per hour per rectangle).

## Sprat 1-ringer mean length from IBTS 1st Q 2007



Figure 8.3.2. SPRAT. Mean length (mm) of age group 1 in the IBTS (February) 2007 in the North Sea and Division IIIa.


Figure 8.3.3. North Sea Sprat. Acoustic survey 2006. Abundance (upper figure in italics, in million individuals) and biomass (lower figure in bold, in kt) per statistical rectangle.


Figure 8.6.1 North Sea sprat. Observed IBTS survey indices for the period 1984 - 2007. Upper figure: recruits (age 1). Lower figure: adults (age 2+).


Figure 8.6.2 North Sea sprat. Correlation between age 1 IBTS index and age $2+$ index in the following year. The correlation between the parameters is weak ( $\mathrm{R}^{2}=0.05$ ).


Figure 8.6.3 North Sea sprat. Normalised cumulative-distribution of the per-haul contribution to the IBTS q1 sprat age 1 survey index. The 300-450 individual-haul contributions to the IBTS index in each year are sorted by size and then aggregated to calculate a cumulative-distribution. The plot shows only the contributions for the 20 largest hauls. Numbers on each line indicate the year for the survey. Similar trends are seen for the IBTS q1 age 2+ index.


Figure 8.6.4 North Sea sprat. Log of the ratio between the age 1 IBTS index and the age 2 index in the subsequent year.


Figure 8.6.5 North Sea sprat. Log of the ratio between landings in two subsequent years from the same year class. The years indicated are landing years and the first in the pair for log landing ratios.


Figure 8.7.1 North Sea sprat. Fit of the CSA model (as described by the sum of squares of the fit error, SSQ) and estimated 2007 total stock biomass for a range of values of the catchability ratio s $=\mathbf{q}_{1} / \mathbf{q}_{2+}$. Note that the estimated 2007 biomass is plotted on a logarithmic scale. Error bars represent a $90 \%$ confidence interval in the biomass, estimated using the non-parametric bootstrapping algorithm in the CSA software. A minima in the $s$ vs SSQ plot occurs at approximately $\boldsymbol{s}=1.075$


Figure 8.7.2 North Sea sprat. Comparison of the IBTS q1 indices (points with thin line) with the fit obtained from the CSA model (thick line). Upper figure: age 1 (recruits) group. Lower figure: age 2+ (mature) group.


Figure 8.7.3 North Sea sprat. Biomass and associated confidence intervals for the time period 1996-2007 as estimated by CSA for a catchability ratio of 1.075 and a natural mortality of 0.75.


Figure 8.7.4 North Sea sprat. Effective fishing mortality ( $\mathrm{F}^{*}$ ) as estimated by the CSA model for a natural mortality of 0.75 and a catchability ratio of $\boldsymbol{s}=\mathbf{1 . 0 7 5}$.


Figure 8.7.5 North Sea sprat. Retrospective analysis of biomass using the CSA model with a natural mortality of 0.75 and catchability ratio, $s=1.075$.


Figure 8.8.1 North Sea sprat. IBTS q1 indices vs total catch (1984-2006). A fitted regression line results in an $R^{2}$ coefficient of 0.47 . The dotted line indicates the IBTS q1 index from 2007 (2688) and gives an estimated 2007 catch of 196 kt. Labels on the plot indicate the year in which the catch was taken.


Figure 8.8.2 North Sea sprat. 2008 sprat biomass (normalised by the estimated 2007 biomass) forecast as a function of the 2007 sprat catch. The heavy central line represents the median estimate, with the lighter lines encompassing the $\mathbf{5 0 \%}$ confidence interval and the lightest outlines covering the $\mathbf{9 0 \%}$ interval. The point to the furthest left-hand side of the figure shows the estimated 2007 with $\mathbf{9 0 \%}$ error bars. The forecast was prepared based on data from a CSA model run with $\mathbf{s = 1 . 0 7 5}$ and a natural mortality of 0.75 .


Figure 8.8.3 North Sea sprat. 2008 forecast biomass as a function of 2007 catch for the catchability ratio, s , of 0.7 . All other parameters, including the axes of the figure, are the same as for Figure 8.8.2.

## 9 Sprat in Division IIIa

### 9.1 The Fishery

### 9.1.1 ACFM advice applicable for 2006 and 2007

The ACFM advice on sprat management is that exploitation of sprat will be limited by the restrictions imposed on fisheries for juvenile herring. This is a result of sprat being fished mainly together with juvenile herring. The sprat fishery is controlled by a herring by-catch quota as well as by-catch percentage limits. No ACFM advice on sprat TAC has been given in recent years. The sprat TAC for 2006 was 52000 t , with a by-catch quota of herring of 20 528 t for the EU fleet. For 2007 the TAC is set at 52000 t and the by-catch of herring at 15 396t.

### 9.1.2 Landings

The total landings almost doubled from 2004 to 2005 but decreased in 2006 to 12570 t, the lowest landings since 1993 (Table 9.1.1). The table present the landings from 1996 onwards. Due to the implementation of the new Danish monitoring scheme, the data from 1996 and onwards are considered reliable in this context.

The reduction in landings in 2006 were mainly in the Danish fishery and seen in both Skagerrak and Kattegat. The Norwegian and Swedish landings include the coastal and fjord fisheries. The data prior to 1996 can be found in the HAWG report from 2006 (ICES 2006/ACFM:20).

In general, there were sprat landings in all quarters (Table 9.1.2). In the first quarter, most of the landings were from Kattegat. Later in the year the landings are mainly reported from the outer Skagerrak (Figures 8.1.1-8.1.2). In 2006 more than $65 \%$ of the total landings were taken in the $1^{\text {st }}$ quarter. In the Norwegian fishery minor landings were taken in the $1^{\text {st }}$ and $4^{\text {th }}$ quarter.

The Danish monitoring scheme for management purposes for species composition in the landings of the Danish small-meshed fisheries has worked well in 2006. A total of 106 samples were collected from all small meshed landings taken in Division IIIa by Danish vessels in 2006. The sampling figure for 2005 was 315 samples. The decrease in number of samples is caused by a significant decrease in total landings for the area. The total landings from the Danish small mesh fishery in 2006 were $20,600 \mathrm{t}$ (all species) compared to $56,800 \mathrm{t}$ in 2005.

### 9.1.3 Fleets

Fleets from Denmark, Norway and Sweden carry out the sprat fishery in Division IIIa.
The Danish sprat fishery consists of trawlers using a 16 mm -mesh size cod end and all landings are used for fishmeal and oil production. Some of the sprat landings from Denmark and Sweden are by-catches in the herring fishery using 32 mm mesh-size cod ends.

There is a Swedish fishery directed at sprat with by-catches of herring. There is also a fishery carried out with small purse seiners at the West Coast of Sweden for human consumption.

The Norwegian sprat fishery in Division IIIa is a coastal/fjord purse seine fishery for human consumption.

### 9.2 Biological Composition of the Catch

### 9.2.1 Catches in number and weight-at-age

Total numbers of sprat taken in Div. IIIA in 2006, were the lowest for the period 1996-2006 (Table 9.2.1). Sprat of the 2004-year class (age 2) made more than $50 \%$ of the total numbers and was one of the largest 2-groups. Most of them were taken in the first quarter. In 2005, the same year class represented approximately $90 \%$ of the total number. The overall mean weight of the 2-group in 2006 were smaller than in 2005 (Table 9.2.2).

Denmark provided biological samples from all the quarters while Sweden provided biological samples from three of the quarters. No Norwegian samples were collected. Landings in 2006, for which samples were collected, were raised using a combination of Swedish and Danish samples, without any differentiation in types of fleets. Details on the sampling for biological data per country, area and quarter are shown in Table 9.2.3.

### 9.3 Fishery-independent information

Acoustic estimates of sprat have been available from the ICES co-ordinated Herring Acoustic surveys in Div. IIIa since 1996. At the time of the surveys, sprat has mainly been recorded in the Kattegat (ICES CM 2006/LRC:04).

In 2005 the abundance and total biomass of sprat in the Kattegat was estimated to 4570 million individuals, equivalent to 54,000 tonnes. In the south western part of the Skagerrak the respective figures were about 490 million individuals, equivalent to 5,800 tonnes.

In 2006 sprat was observed in the ICES squares 41G1-G2, 42G1-G2 and 44G0-G1, all in the Kattegat. The abundance and total biomass were estimated to 2242 million individuals, a decrease of more than $50 \%$ compared to the previous year. The biomass was estimated to 34 , 000 t, of which immature fish made $63 \%$.

The IBTS (February) sprat indices for 1984-2006 are presented in Table 9.3.1. The IBTS data are provided by rectangle in Figure 8.3.1 for age groups 1, 2 and 3+, and the mean length ( mm ) of 1-ringer sprat in Figure 8.3.2. The indices are calculated as mean no./hr (CPUE) weighted by area where water depths are between 10 and 150 m (ICES 1995/Assess:13). The indices were revised in 2002 (ICES 2002/ACFM:12) based on an agreement in the IBTS WG in 1999, where it was decided to calculate the sprat index as an area weighted mean over means by rectangles for the IIIa (ICES 1999/D:2). The old time-series of IBTS indices (from 1984-2001) is shown in ICES 2001/ACFM:10.

The 2007 total IBTS index for 2007 declined compared to 2006-index which was very high, and the highest for the time series. This year the total index was one of the lowest for the period. The very strong 2004-year class, representing $77 \%$ of the total index last year, was not showing up as an extraordinary year class as 3 years old. The extraordinarily high index for 2group sprat last year was based on one single haul, which gives rise to doubts over the validity of this index-value.

### 9.4 Mean weight-at-age

Mean weights-at-age (g) in the catches are presented, by quarter, in Table 9.2.2. The table includes mean weights-at-age for 1996-2005 for comparison. These have been very variable over time, but whether this is due to actual variation in mean weight or difficulties in ageing of sprat is uncertain.

### 9.5 Recruitment

For this stock the IBTS index for 1-group sprat in the first quarter is considered the most suitable recruitment index. The 1-group index for 2006 is at the level of the average for the time-series (1984-2006). The procedure for the survey did not differ from previous years. However, the index does not fully reflect strong and weak cohorts in sprat seen in the catch. This was also expressed in previous working group report (ICES 1998 ACFM :14). This can still be linked to difficulties in age determination

### 9.6 State of the Stock

No assessments of the sprat stock in Division IIIa have been presented since 1985 and this year is no exception. A Schaefer model was fit to the data in 1999 (ICES 1999/ACFM:12) but that attempt was not successful and was subsequently abandoned. In 2003 and 2004 the Working Group agreed to explore the data for sprat in Division IIIa by means of Catch-Survey Analysis (CSA) as performed for sprat in the North Sea (ICES 2003/ACFM:17). This was redone last year with the time series 1994-2005. The mean weights used were the same as for the North Sea, except for the three most recent years, where mean weight at age was available from the IBTS database. The attempt was not successful, suggesting scaling problems and input-data problems, which have not yet been solved by the working group. No new input-data was available for the HAWG, thus no exploratory CSA-runs were made this year.

The signal in the IBTS (February)-index for 2007 indicates a decrease in the sprat stock from last year and appears to be one of the lowest for the time-series 1984-2006.

### 9.7 Projection of Catch and Stock

There is no relationship between the IBTS (February) index (no./h) and the total catch in the same year and the index is not considered useful for management of sprat in Division IIIa.

### 9.8 Reference Points

There are no reference points defined for this stock.

### 9.9 Management Considerations

Sprat in Division IIIa is short-lived with large inter-annual fluctuations in stock biomass. The natural inter-annual variability in stock abundance, mainly driven by recruitment variability, is high and does not appear to be strongly influenced by the observed levels of fishing effort.

The sprat has mainly been fished together with herring, except for 1994 and 1995 when a directed sprat fishery was carried out with low by-catches of herring. The human consumption fishery takes only a minor proportion of the total catch. With the current management regime, where there is a by-catch ceiling limitation of herring as well as by-catch percentage limits, the sprat fishery is controlled by these factors. In the last years the sprat fishery was limited by quota restriction on sprat and not by by-catch restrictions on herring. The same situation may occur in 2007.

Attempts to assess this stock have demonstrated the need for:

- Development of a suitable biomass index
- Improvement of the ageing techniques

Effort should be allocated into the development of a more suitable method for projection of catch and stock. There is also a need for better knowledge of spawning seasons and possible recruitment from the North Sea stock.

Table 9.1.1 Division IIIa sprat. Landings in ('000 t) 1996-2006.
(Data provided by Working Group members). These figures do not in all cases correspond to the official statistics and cannot be used for management purposes.

| Year | Skagerrak |  |  |  | Kattegat |  |  | $\begin{gathered} \hline \text { Div. IIIa } \\ \text { total } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Denmark | Sweden | Norway | Total | Denmark | Sweden | Total |  |
| 1996 | 7.0 | 3.5 | 1.0 | 11.5 | 3.4 | 3.1 | 6.5 | 18.0 |
| 1997 | 7.0 | 3.1 | 0.4 | 10.5 | 4.6 | 0.7 | 5.3 | 15.8 |
| 1998 | 3.9 | 5.2 | 1.0 | 10.1 | 7.3 | 1.0 | 8.3 | 18.4 |
| 1999 | 6.8 | 6.4 | 0.2 | 13.4 | 10.4 | 2.9 | 13.3 | 26.7 |
| 2000 | 5.1 | 4.3 | 0.9 | 10.3 | 7.7 | 2.1 | 9.8 | 20.1 |
| 2001 | 5.2 | 4.5 | 1.4 | 11.2 | 14.9 | 3.0 | 18.0 | 29.1 |
| 2002 | 3.5 | 2.8 | 0.0 | 6.3 | 9.9 | 1.4 | 11.4 | 17.7 |
| 2003 | 2.3 | 2.4 | 0.8 | 5.6 | 7.9 | 3.1 | 10.9 | 16.5 |
| 2004 | 6.2 | 4.5 | 1.1 | 11.8 | 8.2 | 2.0 | 10.2 | 22.0 |
| 2005 | 12.1 | 5.7 | $0.7{ }^{\circ}$ | 18.5 | 19.8 | 2.1 | 21.8 | 40.3 |
| 2006 | 1.2 | 2.8 | 0.3 | 4.3 | 6.6 | 1.6 | 8.2 | 12.5 |

Table 9.1.2. Division Illa sprat. Landings of sprat ('OOO t) by quarter by countries, 1996-2006.
(Data provided by the Working Group members)

|  | Quarter | Denmark | Norway | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 1 | 5.6 | + | 4.2 | 9.8 |
|  | 2 | 3.4 |  | $0.2^{\text {F }}$ | 3.6 |
|  | 3 | + | 0.4 | + | 0.4 |
|  | 4 | 1.4 | 0.6 | $2.2{ }^{\text {F }}$ | 4.2 |
|  | Total | 10.4 | 1.0 | 6.6 | 18.0 |
| 1997 | 1 | 0.7 | - | 0.3 | 1.0 |
|  | 2 | 0.4 | - | 1.2 | 1.6 |
|  | 3 | 2.3 | - | 0.1 | 2.4 |
|  | 4 | 8.2 | 0.4 | 2.2 | 10.8 |
|  | Total | 11.6 | 0.4 | $3.8{ }^{\prime}$ | 15.8 |
| 1998 | 1 | 4.0 | 0.1 | 0.1 | 4.2 |
|  | 2 | 0.9 |  | + | 0.9 |
|  | 3 | 1.1 | 0.3 | 0.4 | 1.8 |
|  | 4 | 5.4 | 0.7 | 5.7 | 11.7 |
|  | Total | 11.4 | 1.1 | $6.1{ }^{\prime}$ | 18.6 |
| 1999 | 1 | 3.5 | 0.0 | $4.0{ }^{\circ}$ | 7.5 |
|  | 2 | 0.1 |  | $0.2^{\prime \prime}$ | 0.3 |
|  | 3 | 7.4 | 0.1 | $1.9{ }^{\text {r }}$ | 9.4 |
|  | 4 | 6.2 | 0.1 | $3.3{ }^{\prime}$ | 9.6 |
|  | Total | 17.2 | 0.2 | $9.3{ }^{\prime}$ | 26.7 |
| 2000 | 1 | 4.1 | 0.1 | 2.3 | 6.5 |
|  | 2 | 0.0 |  | $1.9{ }^{\text {r }}$ | 1.9 |
|  | 3 | 4.8 | 0.1 | $0.0^{\prime}$ | 4.9 |
|  | 4 | 3.8 | 0.7 | $2.3{ }^{\prime}$ | 6.8 |
|  | Total | 12.7 | $0.9{ }^{\prime}$ | 6.4 | 20.0 |
| 2001 | 1 | 2.5 |  | 2.6 | 5.2 |
|  | 2 | 6.6 |  | $0.1{ }^{\text {r }}$ | 6.7 |
|  | 3 | 10.2 |  | $0.1^{\text {r }}$ | 10.2 |
|  | 4 | 0.9 | 1.4 | $4.8{ }^{\prime \prime}$ | 7.1 |
|  | Total | 20.2 | 1.4 | $7.6^{\prime \prime}$ | 29.1 |
| 2002 | 1 | 3.8 | 0.0 | 1.4 | 5.2 |
|  | 2 | 2.1 |  | $0.4{ }^{\text {F }}$ | 2.4 |
|  | 3 | 5.9 | 0.0 | $0.1^{\prime}$ | 6.0 |
|  | 4 | 1.7 | 0.0 | $2.4{ }^{\text {r }}$ | 4.1 |
|  | Total | 13.4 | 0.0 | 4.3 | 17.7 |
| 2003 | 1 | 3.5 | 0.1 | 1.7 | 5.3 |
|  | 2 | 0.6 |  | 0.8 | 1.4 |
|  | 3 | 1.0 |  | 0.7 | 1.7 |
|  | 4 | 5.0 | 0.8 | 2.3 | 8.1 |
|  | Total | 10.2 | 0.8 | 5.5 | 16.5 |
| 2004 | 1 | 3.1 | 0.0 | 1.4 | 4.5 |
|  | 2 | 0.6 |  | 0.9 | 1.5 |
|  | 3 | 3.7 |  | 0.4 | 4.1 |
|  | 4 | 6.9 | 1.1 | 3.8 | 11.9 |
|  | Total | 14.4 | 1.1 | 6.5 | 22.0 |
| 2005 | 1 | 6.5 |  | 1.7 | 8.1 |
|  | 2 | 4.6 |  | $0.1^{\prime}$ | 4.7 |
|  | 3 | 18.6 | 0.7 | $0.8{ }^{\prime \prime}$ | 20.1 |
|  | 4 | 2.1 |  | $5.2{ }^{\prime}$ | 7.3 |
|  | Total | 31.9 | 0.7 | $7.7^{\prime}$ | 40.3 |
| 2006 | 1 | 5.4 | 0.2 | 2.7 | 8.3 |
|  | 2 | 0.2 |  | 0.2 | 0.3 |
|  | 3 | 1.3 |  | 0.1 | 1.4 |
|  | 4 | 0.9 | 0.1 | 1.5 | 2.5 |
|  | Total | 7.8 | 0.3 | 4.4 | 12.5 |

+ Catch record, but amount not precisely known.
${ }^{1}$ Preliminary figures

Table 9.2.1 Division IIla sprat. Landed numbers (millions) of sprat by age groups in 1996-2006.

|  | Quarter |  |  | Age |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0 | 1 | 2 | 3 | 4 | 5+ |  |
| 1996 | 1 |  | 288.42 | 546.53 | 62.11 | 15.65 | 5.07 | 917.78 |
|  | 2 |  | 0.89 | 414.10 | 42.76 | 0.71 | 0.06 | 458.51 |
|  | 3 |  | 0.34 | 1.81 | 0.30 | 0.02 |  | 2.47 |
|  | 4 |  | 31.19 | 165.65 | 27.34 | 2.03 |  | 226.21 |
|  | Total |  | 320.84 | $1128.08{ }^{\text {² }}$ | $132.51{ }^{\circ}$ | 18.41 | 5.13 | 1,604.97 |
| 1997 | 1 |  |  | 3.43 | 18.31 | 20.60 | 4.59 | 46.94 |
|  | 2 |  | 1.00 | 2.76 | 19.56 | 1.51 | 0.25 | 25.07 |
|  | 3 | 4.35 | 209.25 | 9.51 | 1.92 | 6.24 |  | 231.26 |
|  | 4 | 32.39 | 644.28 | 58.31 | 7.16 | 28.02 |  | 770.16 |
|  | Total | 36.74 | 854.53 | 74.01 | 46.95 | 56.37 | 4.84 | 1,073.43 |
| 1998 | 1 |  | 14.91 | 103.38 | 94.00 | 76.99 | 6.34 | 295.61 |
|  | 2 |  | 3.24 | 21.49 | 20.59 | 16.63 | 1.33 | 63.28 |
|  | 3 | 53.62 | 26.03 | 41.84 | 5.65 | 0.74 |  | 127.88 |
|  | 4 | 192.13 | 253.98 | 226.55 | 53.14 | 29.80 |  | 755.61 |
|  | Total | 245.75 | 298.16 | 393.25 | 173.38 | 124.17 | 7.67 | 1,242.38 |
| 1999 | 1 | 0.0 | 560.5 | 158.0 | 151.2 | 77.4 | 6.8 | 953.9 |
|  | 2 |  | 32.8 | 1.6 | 1.7 | 1.1 | 0.3 | 37.6 |
|  | 3 | 9.6 | 741.7 | 46.7 | 6.3 | 5.9 |  | 810.0 |
|  | 4 | 8.5 | 645.4 | 20.5 | 6.8 | 0.6 | 0.3 | 682.1 |
|  | Total | 18.0 | 1,980.4 | 226.8 | 166.0 | 85.0 | 7.4 | 2,483.6 |
| 2000 | 1 |  | 116.6 | 384.3 | 40.3 | 7.3 | 1.6 | 550.0 |
|  | 2 |  | 17.3 | 127.4 | 11.2 |  |  | 155.9 |
|  | 3 | 2.1 | 223.3 | 51.4 | 12.2 |  |  | 289.1 |
|  | 4 | 18.0 | 277.6 | 81.4 | 13.1 | 0.8 |  | 390.9 |
|  | Total | 20.2 | 634.8 | 644.6 | 76.8 | 8.1 | 1.6 | 1,386.0 |
| 2001 | 1 | 0.0 | 342.6 | 173.0 | 73.3 | 10.0 | 1.6 | 600.4 |
|  | 2 | 0.0 | 1746.4 | 13.4 | 0.4 | 0.0 | 0.0 | 1,760.2 |
|  | 3 | 5.7 | 924.1 | 31.7 | 0.0 | 0.0 | 0.0 | 961.5 |
|  | 4 | 22.9 | 488.1 | 39.1 | 18.5 | 1.5 | 0.5 | 570.6 |
|  | Total | 28.6 | 3,501.2 | 257.2 | 92.2 | 11.5 | 2.1 | 3,892.8 |
| 2002 | 1 | 0.0 | 63.8 | 323.2 | 38.5 | 24.7 | 2.4 | 452.6 |
|  | 2 | 0.0 | 185.5 | 63.2 | 4.8 | 1.0 | 0.0 | 254.5 |
|  | 3 | 1.3 | 326.2 | 102.0 | 23.9 | 6.6 | 0.6 | 460.5 |
|  | 4 | 21.3 | 205.4 | 45.9 | 10.6 | 5.9 | 0.4 | 289.6 |
|  | Total | 22.5 | 780.9 | 534.3 | 77.9 | 38.2 | 3.4 | 1,457.2 |
| 2003 | 1 | 0.0 | 17.5 | 221.4 | 100.7 | 17.6 | 4.3 | 361.5 |
|  | 2 | 0.0 | 2.6 | 49.8 | 24.0 | 5.5 | 2.1 | 84.1 |
|  | 3 | 192.7 | 10.9 | 31.6 | 5.4 | 2.7 | 0.0 | 243.3 |
|  | 4 | 321.6 | 131.7 | 100.6 | 42.5 | 3.4 | 2.3 | 602.2 |
|  | Total | 514.3 | 162.7 | 403.4 | 172.6 | 29.2 | 8.8 | 1,291.1 |
| 2004 | 1 |  | 539.6 | 39.3 | 47.2 | 20.7 | 8.0 | 654.8 |
|  | 2 |  | 36.7 | 22.3 | 44.9 | 11.8 | 1.1 | 116.8 |
|  | 3 | 10.0 | 254.4 | 19.4 | 4.1 | 2.4 |  | 290.3 |
|  | 4 | 874.0 | 366.8 | 33.0 | 24.9 | 3.4 | 0.3 | 1,302.3 |
|  | Total | 883.9 | 1,197.5 | 113.9 | 121.1 | 38.3 | 9.3 | 2,364.2 |
| 2005 | 1 | 0.0 | 1609.1 | 185.6 | 25.5 | 17.4 | 5.1 | 1,842.7 |
|  | 2 | 0.0 | 827.1 | 19.2 | 0.6 | 0.0 | 0.0 | 846.9 |
|  | 3 | 1.8 | 1557.0 | 91.3 | 9.9 | 12.9 | 0.0 | 1,672.9 |
|  | 4 | 11.5 | 447.4 | 60.5 | 7.3 | 4.0 | 0.7 | 531.3 |
|  | Total | 13.4 | 4,440.6 | 356.6 | 43.3 | 34.2 | 5.8 | 4,893.9 |
| 2006 | 1 | 0.0 | 219.8 | 433.3 | 93.7 | 16.6 | 10.3 | 773.7 |
|  | 2 | 0.0 | 7.5 | 17.8 | 1.6 | 0.3 | 0.0 | 27.2 |
|  | 3 | 0.0 | 9.4 | 55.8 | 13.7 | 2.8 | 1.3 | 83.1 |
|  | 4 | 4.0 | 38.5 | 71.6 | 18.4 | 0.9 | 0.7 | 134.0 |
|  | Total | 4.0 | 275.2 | 578.5 | 127.4 | 20.6 | 12.3 | 1,018.0 |

Table 9.2.2. Division IIIa Sprat. Quarterly mean weight-at-age ( g ) in the landings.
(1998-2006 Danish and Swedish data, 1996-1997 Danish data, 2006)

| Year | Age |  | 2 | 3 | 4 | 5+ | SOPCorrected landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarter | 0 | 1 |  |  |  |  |  |
| 1996 |  | 9.2 | 10.6 | 14.2 | 17.4 | 17.7 | 9,724 |
|  |  | 8.6 | 12.5 | 15.1 | 17.4 | 17.0 | 5,847 |
|  |  | 4.2 | 10.9 | 15.5 | 21.0 |  | 26 |
|  |  | 4.2 | 10.9 | 15.5 | 21.0 |  | 2,403 |
| Weighted mean |  | 8.7 | 7.6 | 14.8 | 19.6 | 17.7 | 18,000 |
| 1997 |  |  | 17.3 | 18.6 | 21.8 | 26.0 | 968 |
|  |  | 8.3 | 17.6 | 20.0 | 22.1 | 31.0 | 489 |
|  | 4.1 | 13.6 | 17.2 | 21.1 |  |  | 3,062 |
|  | 4.7 | 14.7 | 17.5 |  | 19.5 |  | 11,176 |
| Weighted mean | 4.6 | 14.4 | 17.5 | 19.6 | 20.4 | 26.3 | 15,696 |
| 1998 |  | 6.6 | 14.0 | 18.0 | 19.0 | 21.3 | 4,828 |
|  |  | 6.6 | 13.9 | 17.8 | 18.7 | 21.0 | 1,027 |
|  | 4.6 | 17.7 | 20.7 | 22.1 | 24.7 |  | 1,718 |
|  | 4.8 | 17.5 | 20.4 | 22.5 | 27.5 |  | 11,998 |
| Weighted mean | 4.8 | 16.9 | 18.5 | 19.6 | 21.2 | 21.2 | 19,570 |
| 1999 1 1 |  | 4.6 | 6.4 | 17.3 | 13.4 | 13.1 | 7,319 |
|  |  | 5.3 | 17.1 | 18.6 | 22.2 | 17.8 | 264 |
|  | 3.0 | 11.4 | 12.6 | 16.8 | 18.3 |  | 9,257 |
|  | 4.8 | 13.9 | 17.6 | 20.8 | 21.2 | 23.5 | 9,521 |
| Weighted mean | 3.8 | 10.2 | 8.8 | 17.4 | 13.9 | 13.7 | 26,361 |
| 2000 |  | 5.3 | 13.1 | 15.3 | 20.7 | 22.7 | 6,438 |
|  |  | 5.2 | 12.8 | 14.1 |  |  | 1,873 |
|  | 4.3 | 16.6 | 18.0 | 21.9 |  |  | 4,897 |
| 4 | 7.0 | 16.9 | 19.9 | 22.1 | 24.6 |  | 6,742 |
| Weighted mean | 6.7 | 14.3 | 14.3 | 17.3 | 21.1 | 22.7 | 19,949 |
| $2001 \begin{array}{ll}1 \\ & 2 \\ & 3 \\ & 4\end{array}$ |  | 3.8 | 14.3 | 16.2 | 17.8 | 17.3 | 5,168 |
|  |  | 3.7 | 6.5 | 21.0 |  |  | 6,598 |
|  | 5.3 | 10.5 | 12.1 |  | 13.0 |  | 10,114 |
|  | 5.1 | 12.0 | 19.7 | 22.6 | 19.3 | 25.6 | 7,200 |
| Weighted mean | 5.1 | 6.7 | 14.5 | 17.5 | 18.0 | 19.2 | 29,079 |
| 2002 1 4 |  | 5.7 | 12.7 | 17.3 | 19.3 | 20.6 | 5,411 |
|  |  | 7.9 | 13.7 | 16.0 | 17.0 |  | 2,175 |
|  | 8.0 | 12.4 | 15.1 | 18.1 | 17.0 | 17.0 | 5,900 |
|  | 5.7 | 15.6 | 18.2 | 21.6 | 21.5 | 22.0 | 4,278 |
| Weighted mean | 5.8 | 11.6 | 13.7 | 18.1 | 19.2 | 20.1 | 17,763 |
| 2003 (r 1 |  | 6.0 | 14.1 | 16.2 | 18.9 | 23.8 | 5,293 |
|  |  | 5.0 | 16.0 | 17.6 | 21.6 | 22.8 | 1,401 |
|  | 4.0 | 12.0 | 19.0 | 19.0 | 21.0 |  | 1,661 |
|  | 8.9 | 16.4 | 21.1 | 21.7 | 25.2 | 24.3 | 8,211 |
| Weighted mean | 7.1 | 14.8 | 16.5 | 17.8 | 20.3 | 23.7 | 16,565 |
| 2004 |  | 4.6 | 14.6 | 17.8 | 17.3 | 17.3 | 4,392 |
|  |  | 7.0 | 13.6 | 16.7 | 17.0 | 19.5 | 1,532 |
|  | 3.0 | 14.1 | 16.7 | 20.0 | 21.4 |  | 4,075 |
| 4 | 3.5 | 16.8 | 19.9 | 22.2 | 20.9 | 28.0 | 10,508 |
| Weighted mean | 3.5 | 10.4 | 16.3 | 18.4 | 17.8 | 17.9 | 20,508 |
| 2005 |  | 3.0 | 14.6 | 16.3 | 20.3 | 21.1 | 8,149 |
|  |  | 5.4 | 11.7 | 26.8 | 0.0 |  | 4,723 |
|  | 2.9 | 11.9 | 14.6 | 15.4 | 11.0 |  | 20,130 |
| 4 | 3.3 | 13.1 | 19.1 | 20.1 | 21.1 | 23.1 | 7,300 |
| Weighted mean | 5.0 | 7.6 | 15.4 | 17.1 | 17.2 | 21.5 | 40,301 |
| 2006 |  | 5.0 | 12.2 | 15.4 | 15.2 | 18.5 | 8,279 |
|  |  | 7.0 | 13.3 | 16.3 | 22.0 |  | 324 |
|  |  | 11.2 | 17.4 | 20.3 | 18.6 | 22.8 | 1,440 |
| 4 | 4.3 | 16.1 | 19.6 | 21.4 | 23.8 | 26.6 | 2,464 |
| Weighted mean | 4.3 | 6.8 | 13.6 | 16.8 | 16.1 | $19.4{ }^{\prime}$ | 12,507 |

Table 9.2.3 Division IIIa sprat. Sampling commercial landings for biological samples in 2006.

| Country | Quarter | Landings <br> (tonnes) | No. <br> samples | No. <br> meas. | No. <br> aged |
| :--- | :---: | ---: | ---: | ---: | ---: |
| Denmark | 1 | 5,428 | 31 | 3,132 | 3,024 |
|  | 2 | 168 | 1 | 100 | 98 |
|  | 3 | 1,343 | 13 | 1,319 | 1,319 |
|  | 4 | 880 | 7 | 585 | 577 |
| Norway | Total | 7,819 | 52 | 5,136 | 5,018 |
|  | 1 | 174 | 0 | 0 | 0 |
|  | 2 | - |  |  |  |
| Sweden | 3 | - |  |  |  |
|  | 4 | 129 | 0 | 0 | 0 |
|  | 1 | 2,677 | 9 | 662 | 662 |
|  | 2 | 156 | 1 | 76 | 76 |
| Denal | 303 | 0 | 0 | 0 |  |
| Norwark | 3 | 97 |  |  | 0 |
| Sweden | 4 | 1,455 | 11 | 697 | 697 |
|  |  | 4,385 | 21 | 1,435 | 1,435 |
|  |  | 7,819 | 52 | 5,136 | 5,018 |

Table 9.3.1. Division IIIa sprat. IBTS(February) indices of sprat per age group 1984-2006. (Mean number per hour per rectangle weighted by area. Only hauls taken in depths of 10-150 m are included).

| Year | No Rect | No hauls | Age Group |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 1 | 2 | 3 | 4 | $5+$ | Total |
| 1984 | 15 | 38 | 5675.45 | 868.88 | 205.10 | 79.08 | $63.5{ }^{\prime \prime}$ | 6,892.08 |
| 1985 | 14 | 38 | 2157.76 | 2347.02 | 392.78 | 139.74 | $51.24{ }^{\prime}$ | 5,088.54 |
| 1986 | 15 | 38 | 628.64 | 1979.24 | 2034.98 | 144.19 | $37.53{ }^{\prime \prime}$ | 4,824.58 |
| 1987 | 16 | 38 | 2735.92 | 2845.93 | 3003.22 | 2582.24 | $156.64{ }^{\prime \prime}$ | 11,323.95 |
| 1988 | 13 | 38 | 914.47 | 5262.55 | 1485.07 | 2088.05 | $453.13^{\prime \prime}$ | 10,203.26 |
| 1989 | 14 | 38 | 413.94 | 911.28 | 988.95 | 554.53 | $135.79{ }^{\prime \prime}$ | 3,004.48 |
| 1990 | 15 | 38 | 481.02 | 223.89 | 64.93 | 61.11 | 45.69" | 876.65 |
| 1991 | 14 | 38 | 492.50 | 726.82 | 698.11 | 128.36 | $375.44^{\prime \prime}$ | 2,421.23 |
| 1992 | 16 | 38 | 5993.64 | 598.71 | 263.97 | 202.90 | $76.04{ }^{\text {r }}$ | 7,135.25 |
| 1993 | 16 | 38 | 1589.92 | 4168.61 | 907.43 | 199.32 | $239.64{ }^{\prime \prime}$ | 7,104.92 |
| 1994 | 16 | 38 | 1788.86 | 715.84 | 1050.87 | 312.65 | $70.11{ }^{\text {F }}$ | 3,938.32 |
| 1995 | 17 | 38 | 2204.07 | 1769.53 | 35.19 | 44.96 | 4.23 " | 4,057.98 |
| 1996 | 15 | 38 | 199.30 | 5515.42 | 692.78 | 111.98 | $173.75^{\prime \prime}$ | 6,693.23 |
| 1997 | 16 | 41 | 232.65 | 391.23 | 1239.13 | 139.14 | 134.51 ${ }^{\prime \prime}$ | 2,136.67 |
| 1998 | 15 | 39 | 72.25 | 1585.22 | 619.76 | 1617.71 | 521.52 ${ }^{\prime \prime}$ | 4,416.46 |
| 1999 | 16 | 42 | 4534.96 | 355.24 | 249.86 | 44.25 | $313.52^{\prime \prime}$ | 5,497.83 |
| 2000 | 16 | 41 | 292.32 | 737.80 | 59.69 | 51.79 | $23.21^{\prime \prime}$ | 1,164.80 |
| 2001 | 16 | 42 | 6539.48 | 1144.34 | 676.71 | 92.37 | $45.87{ }^{\prime \prime}$ | 8,498.77 |
| 2002 | 16 | 42 | 1180.52 | 1035.71 | 89.96 | 58.85 | 12.93 | 2,241.90 |
| 2003 | 17 | 46 | 462.64 | 1247.49 | 1172.13 | 382.29 | 123.17 | 3,387.72 |
| 2004 | 16 | 41 | 402.87 | 49.00 | 156.62 | 86.57 | $27.48{ }^{*}$ | 722.54 |
| 2005 | 17 | 50 | 3314.17 | 1563.16 | 470.84 | 837.09 | 538.37 | 6,722.82 |
| 2006 | 17 |  | 1323.59 | 11855.76 | 1753.92 | 299.05 | 159.23 | 15,391.55 |
| 2007 |  |  | 774.11 | 306.63 | 250.81 | 42.08 | 13.74 | 1,387.37 |

## 10 Stocks with insufficient data

Two stocks with very low research intensity were poorly described in previous reports in devoted sections or chapters. These were Clyde herring (section 5.11 in ICES 2005a) and sprat in VIId,e (section 9, in ICES 2005a). The advice on these stocks cannot be improved at present. In this section only the times series are maintained. For most recent advice refer to the appropriate sections in last year's HAWG report (ICES 2005a).

There was zero sampling of the catch in 2006 for both Clyde herring and sprat in VIId,e. The catch of Clyde herring in 2006 was low (Table 10.1) as was the catch of sprat in VIId,e (Table 10.2).

Table 10.1
Herring from the Firth of Clyde. Catch in tonnes by country, 1955-2006. Spring and autumn-spawners combined.

| Year | 1955 | 1956 | 1957 | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| All Catches |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 4050 | 4848 | 5915 | 4926 | 10530 | 15680 | 10848 | 3989 | 7073 | 14509 | 15096 | 9807 | 7929 | 9433 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Year | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 |  |
| All Catches |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 10594 | 7763 | 4088 | 4226 | 4715 | 4061 | 3664 | 4139 | 4847 | 3862 | 1951 | 2081 | 2135 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Year | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| Scotland | 2506 | 2530 | 2991 | 3001 | 3395 | 2895 | 1568 | 2135 | 2184 | 713 | 929 | 852 | 608 | 392 |
| Other UK | - | 273 | 247 | 22 | - | - | - | - | - | - | - | 1 | - | 194 |
| Unallocated ${ }^{1}$ | 262 | 293 | 224 | 433 | 576 | 278 | 110 | 208 | 75 | 18 | - | - | 2 | - |
| Discards | 1253 | 1265 | $2308{ }^{3}$ | $1344{ }^{3}$ | $679{ }^{3}$ | $439{ }^{4}$ | $245{ }^{4}$ | - ${ }^{2}$ | - ${ }^{2}$ | - ${ }^{2}$ | $-{ }^{2}$ | $-{ }^{2}$ | $-^{2}$ | $-{ }^{2}$ |
| Agreed TAC |  |  | 3000 | 3000 | 3100 | 3500 | 3200 | 3200 | 2600 | 2900 | 2300 | 1000 | 1000 | 1000 |
| Total | 4021 | 4361 | 5770 | 4800 | 4650 | 3612 | 1923 | 2343 | 2259 | 731 | 929 | 853 | 608 | 586 |



Table 10.2. Sprat VIId,e. Nominal catches of sprat in VIId,e from 1985-2006

| Country | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  | 15 | 250 | 2529 | 2092 | 608 |  |  |
| France | 14 |  | 23 | 2 | 10 |  |  | 35 |
| Netherlands |  |  |  |  |  |  |  |  |
| UK (Engl.\&Wales) | 3771 | 1163 | 2441 | 2944 | 1319 | 1508 | 2567 | 1790 |
| Total | 3785 | 1178 | 2714 | 5475 | 3421 | 2116 | 2567 | 1825 |
| Country | 1993 | 1994 | 1995 | 1996 | 1997 | 1998* | 1999* | 2000* |
| Denmark |  |  |  |  |  |  |  |  |
| France | 2 | 1 | 0 |  |  |  |  | 18 |
| Netherlands |  |  |  |  |  |  | 1 | 1 |
| UK (Engl.\&Wales) | 1798 | 3177 | 1515 | 1789 | 1621 | 2024 | 3559 | 1692 |
| Total | 1800 | 3178 | 1515 | 1789 | 1621 | 2024 | 3560 | 1711 |
|  |  |  |  |  |  |  |  |  |
| Country | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |  |  |
| Denmark |  |  |  |  |  |  |  |  |
| France |  |  |  |  |  |  |  |  |
| Netherlands |  |  |  |  |  |  |  |  |
| UK (Engl.\&Wales) | 1349 | 1196 | 1377 | 836 | 1635 | 1974 |  |  |
| Total | 1349 | 1196 | 1377 | 836 | 1635 | 1974 |  |  |
| * Preliminary |  |  |  |  |  |  |  |  |

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## 13-22 March 2007

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## Annex 2: Recommendations

HAWG 2007 makes the following recommendations:

| REcommENDATION | Action |
| :--- | :--- |
| HAWG has a list of recommendations on the utility and development of <br> INTERCATCH. See chapter 1.5.5 of the HAWG report for full details. | ICES Secretariat |
| HAWG recommends that all metiers with substantial catch should be <br> sampled (including by-catches in the small meshed fishery). | National labs |
| HAWG recommends that a project and/or ICES study group be set up to <br> consider the management of herring stocks to the west of the British <br> Isles, in light of the findings of WESTHER. The full recommendation <br> is in chapter 1.3 of this report. | ICES RMC, LRC, national labs, <br> the European Commission |
| HAWG recommends that surveys used in stock assessments of herring <br> and sprat, under the auspices of HAWG should qualify under the DCR <br> and should be coordinated through PGHERS. | EU STECF, European <br> Commission, Nationl Labs |
| HAWG recommends that the micro-increment analysis of otoliths (to <br> determine spawning type) for herring in the North Sea be carried out on <br> samples collected during the annual acoustic survey and on the <br> commercial catches. | National Labs |
| HAWG recommends that ICES make available a time series of IBTS <br> anchovy and sardine CPUE (with an assessment of variance) for the <br> next meeting of HAWG in 2008. | ICES |
| HAWG recommends further work to identify the causes and dynamics <br> of the serial poor recruitment of North Sea herring. | ICES WGRP, SGRECVAP |
| HAWG request that PGNAPES makes available for HAWG 2008, a <br> time series of the abundance at age, of North Sea herring in the | ICES PGNAPES |
| Norwegian Ecosystem survey (with variance), also with the associated |  |
| target strengths used to determine those estimates of abundance. |  |$\quad$ ICES DATRAS.

## Terms of Reference for 2008

The Herring Assessment Working Group for the Area South of $62^{\circ} \mathrm{N}$ (HAWG) (Co-Chairs: Tomas Gröhsler, Germany and Emma Hatfield, UK) will meet in Copenhagen, Denmark from 11 - 20 March 2008 to:
a) assess the status of and provide management options (by fleet where possible) for 2009 for:

- the North Sea autumn-spawning herring stock in Division IIIa, Subarea IV, and Division VIId (separately, if possible, for Divisions IVc and VIId). Forecasts should be provided by fleet if possible and taking into account the management plan agreed between the EU and Norway;
- the herring stocks in Division VIa and Sub-area VII;
- the stock of spring-spawning herring in Division IIIa and Subdivisions 22-24 (Western Baltic); Management options for Division IIIa shall be given by fleets taking into account that North Sea herring and Western Baltic herring are taken together in this Division;
b) assess the status of the sprat stocks in Subarea IV and Divisions IIIa and VIId,e;
c) for the stocks mentioned in a) and b) perform the tasks described in C.Res. 2007/2/ACFM01.

HAWG will report by 1 April to the attention of ACFM.

## Quality Handbook ANNEX: HAWG-herringWBSS

Stock specific documentation of standard assessment procedures used by ICES and relevant knowledge of the biology.

| Stock | Western Baltic Spring spawning herring (WBSS) |
| :--- | :--- |
| Working Group: | Herring Assessment Working Group for the Area <br> South of $62^{\circ} \mathrm{N}$ |
| Date: | 21.03 .2007 |
| Authors: | M. Cardinale, J. Dalskov, T. Gröhsler, H. <br> Mosegaard, M. van Deurs, J. Gröger |

## A. General

## A.1. Stock definition and biology

## Stocks

Herring caught in Division IIIa and the eastern North Sea is a mixture of two stocks: North Sea Autumn Spawners (NSAS) and Western Baltic Spring Spawners (WBSS). All springspawning herring in the eastern part of the North Sea (IVa\&b east), Skagerrak (Sub-division 20), Kattegat (Subdivision 21) and the Western Baltic (Subdivisions 22, 23 and 24) are treated as one stock, WBSS. The main spawning area of the WBSS is considered to be Greifswalter Bodden at Rügen (therefore also referred to as the Rügen-herring) (ICES, 1998), whereas NSAS utilizes spawning areas mainly along the British east coast (e.g. Burd, 1978; Zijlstra, 1969). The assessment aslo take into account the few Norwegian Spring Spawners (NSS) caught in IVa north.

A third well defined stock component in ICES Division IIIa is winter-spawning Downs herring from spawning areas in the English Channel (Bridger, 1961). There is some disagreement on whether this stock in fact should be regarded as an extension of the NSAS. The contribution of Downs-herring to the mix-area of Division IIIa is likely to be relatively small (un-published data from otolith readings, DIFFRES), and Downs-herring are therefore included under NSAS for the stock assessment of herring in Division IIIa and Subdivision 2224.

In the Western Baltic almost solely WBSS are being caught (few autumn spawners, however, have been observed). The majority of 2+ ringers, however, migrate out of the area during quarter 2, to feed in Division IIIa and the North Sea, and return in quarter 1 (Biester, 1979; Nielsen et al., 2001; van Deurs and Ramkaer, submitted in December 2006).

In the Kattegat and the eastern Skagerrak, mainly 2+ ringers of the WBSS and 0 to 2-ringers from the NSAS are being caught (ICES, 2004; ICES WD, 2006). The area provides a nursery habitat for juvenile NSAS (also areas in the North Sea works as nursery areas), that assumable have drifted into the area as larvae (Burd, 1978; Heath et al, 1997). 0-1 ringer WBSS mainly uses nursery areas in Subdivision 22-24 and start to occur in the southern Kattegat as 1ringers. The largest concentrations of herring during June/July seems to appear along the southern edge of the Norwegian Trench and in the area to the east of Læsø, in Kattegat (ICES, 2005; ICES, 2006). In $3^{\text {rd }}$ quarter large concentrations of $2+$ ringers of the WBSS are found in
the southern Kattegat and Subdivision 23 as they aggregate for the over-wintering, which mainly takes place in Subdivision 23 (Nielsen et al., 2001; Clausen et al., 2006).

In the eastern North Sea and the western Skagerrak mainly 2+ ringers from WBSS and 1 to 2ringer NSAS are being caught (Clausen et al., 2006). Peak catches of WBSS occur in quarter 3, during which the spawning stock of WBSS feed in these areas (ICES, 2002). According to the herring acoustic survey (ICES, 2006) the largest concentrations of herring in this area occur along the transition zone between the Skagerrak and the North Sea (ICES, 2006). Some $2+$ ringer NSAS are caught in $1^{\text {st }}$ and $4^{\text {th }}$ quarter, since part of the NSAS spawning stock overwinter in the Norwegian trench in this area. (Burd, 1978; Cushing and Bridger, 1966; Clausen et al., 2006).

In historic time several local spring spawning populations in the Skagerrak and the Kattegat has been described (e.g. Ackerfors, 1977; Rosenburg and Palmen, 1982). The largest of these seems to have reached extinction decades ago (ICES, 2004). Local spawning events during spring in a rather large number of fjords on the coast of Skagerrak and Kattegat, and both in Denmark, Sweeden, and Norway are known still to occur regularly (HERGEN, EU project QLRT 200-01370, final report), but have been considered of minor importance for the herring fisheries (ICES, 2001). Recent genetic and morphological studies confirmed that these local spawning areas belong to distinct spawning populations (Bekkevold et al., 2005), and bear witness of a more complex composition of multiple populations than previously assumed. The migration behaviour of these populations is basically unknown and the methods for splitting them from the Rügen-herring in catches are still associated with large uncertainties (HERGEN, EU project QLRT 200-01370, final report). Also on the German coast of the Western Baltic we find more than the spawning grounds of Rügen. E.g. the spring spawning grounds of the Sleich Fjord (Kühlmorgen-Hille, 1983). It is unknown whether herring visiting spawning grounds in the Sleich Fjord belong to the Rügen-herring or should be considered an independent population. However, results presented by Biester (1979) and the population diversity found by Bekkevold et al. (2005) indicates that they too are likely to be genetically distinct from the Rügen-herring.

## Methods for stock separation

Experience within the Herring Assessment Working Group has shown that stock separation procedures based on size distributions often will fail.

The method for separating herring stocks in Norwegian samples, using vertebral counts (vc), as described in former reports of this Working Group (ICES 1991/ Assess:15), assumes that for NSAS, the mean vertebral count is 56.5 and for WBSS 55.8. The fractions of spring spawners (fsp) are estimated from the formula (56.50-v)/(56.5-55.8), where v is the mean vertebral count of the (mixed) sample with the restriction that the proportion should be one if $\mathrm{fsp}>=1$ and zero if $\mathrm{fsp}<=0$. The method is quite sensitive to within-stock variation (e.g. between year classes) in mean vc. The mean vc, of the previous mentioned local springspawners from the Norwegian Skagerrak fjords (it should be emphasised that this is not the Norwegian Spring Spawners alias Atlantic-Scandio Herring), is higher than for the NSAS (Rosenberg and Palmén, 1982; van Deurs, 2005), and will bias fsp estimates if present in the samples. The Norwegian samples used in the stock assessment are from the eastern North Sea. The local Norwegian spring spawners therefore only constitute a problem if they migrate to feeding areas in the eastern North Sea. Inconclusive results from a study of the tag pratsite $A$. simplex in herring, indicates that this may be the case (van Deurs and Ramkaer, submitted in December 2006).

The introduction of otolith microstructure analysis in 1996-97 (Mosegaard and Popp-Madsen, 1996) enables an accurate and precise split between three groups, autumn, winter and springspawners. Today this method is applied for the stock separation in all Danish and Swedish IIIa
samples. However, different populations with similar spawning periods are not resolved with the present level of analysis. Different stock components that are not easily distinguished by their otolith microstructure (OM) are considered to have different mean vertebral counts (vc): E.g. the local Skagerrak winter/spring-spawners: 57 (Rosenberg and Palmén, 1982); Western Baltic Sea: 55.6 - 55.8 (Gröger and Gröhsler, 2001; ICES 1992/H:5). It should, however, be noted that the estimated stock specific mean vc varies somewhat among different studies, and the vc alone is not likely to be a successful tool for distinguishing between separate spring spawning populations in an assessment context .

Comparison between separation methods using frequency distributions of vertebral counts and otolith microstructure showed reasonable correspondence. Using this information the years from 1991 to 1996 was reworked in 2001, applying common splitting keys for all years by using a combination of the vertebral count and otolith microstructure methods (ICES, 2001). From 2001 and onwards, the otolith-based method only has been used for the Division IIIa.

Different methods of identifying herring stocks in the Division IIIa and Subdivisions 22-24 were recently evaluated in an EU CFP study project (EC study 98/026). The study involved several inter-calibration sessions between microstructure readers in the different laboratories involved with the WBSS herring. After the study was finished a close collaboration concerning reader interpretations has been kept between the Danish and Swedish laboratories. Sub-samples of the 2002 and 2003 Danish, Swedish, and German microstructure analyses were double-checked by the same Danish expert reader for consistency in interpretation. The overall impression is an increasingly good agreement among readers.

New molecular genetic approaches for stock separation are being developed within the EUFP5 project HERGEN (EU project QLRT 200-01370, final report). Sampling of spawning aggregations during spring, autumn and winter has been carried out in 2002 and in 2003 in Division IIIa and in the Western Baltic at more than 10 different locations. The results point at a substantial genetic variation between North Sea and Western Baltic herring. As mentioned earlier, significant variation has also been found among spawning populations in Division IIIa and subdivision 22-24, which indicates the presence of multiple distinct spring spawning populations or sub-populations (Bekkevold et al., 2005). However, the substantial overlap in the genetic profiles of these sub-populations results in large uncertainties when attempting to estimate the proportional contribution of the individual spring spawning populations to the mix in Division IIIa.

For Subdivisions 22, 23 and 24 it is assumed that all individuals caught belong to the WBSS. However, after the introduction of OM analysis in 1996/97 it was discovered that in the western Baltic a small percentage of the herring landings might consist of autumn spawning individuals. Before molecular genetic methods became available for Atlantic herring the existence of varying proportions of autumn spawners in Subdivisions 22-24 in different years was considered a potential problem for the assessment, since they were thought to belong to the NSAS. Today the molecular genetic methods have revealed that they are more closely related to the WBSS than to the NSAS (HERGEN, EU project QLRT 200-01370, final report). Therefore, with the present genetic perception in mind, when herring with OM indicating autumn hatch are found in subdivisions 22-24 these are treated as belonging to the WBSS stock.

OM analysis for stock splitting is a relatively time consuming method, furthermore, its potential for making splits, between the recently discovered complexity of different spring spawning populations, is very limited (un-published results, DIFFRES). Time has therefore been put into developing new, and more time efficient methods, for stock splitting. Under the EU-FP5 project HERGEN (EU project QLRT 200-01370, final report) a promising and time effective method based on otolith morphology are being developed. So far this work has showed that individual stocks and local populations display significantly different edge pattern
of lobe formation in the otolith (the work was conducted on the saggitae otolith). The procedure involves photographing the shapes of the otolith edge and subsequent analysis in the photo treatment software Image Pro plus 5.0. However, so far the technique does not provide a way to efficiently split between spring spawning population in the mix-area of IIIa.

## A.2. Fishery

## Fleet definitions

The fleet definitions used since 1998 for the fishery in Division IIIa are:

- Fleet C: directed fishery for herring in which trawlers (with 32 mm minimum mesh size) and purse seiners participate.
- Fleet D: All fisheries in which trawlers (with mesh sizes less than 32 mm ) and small purse seiners, fishing for sprat along the Swedish coast and in the Swedish fjords, participate. For most of the landings taken by this fleet, herring is landed as by-catch.

Danish and Swedish by-catches of herring from the sprat fishery and the Norway pout and blue-whiting fisheries are listed under fleet D.

In SDs 22-24 most of the catches are taken in a directed fishery for herring and some as bycatch in a directed sprat fishery. All landings from SDs22-24 are treated as one fleet.

## Historical German fishing pattern

The overall German fishing pattern has changed in the last few years. Until 2000 the dominant part of of German herring cacthes were caught in the passive fishery by gillnets and trapnets around the Rügen Island. Since 2001 the activities in the trawl fishery increased. Recently the landings by trawl reached a level of more than $50 \%$ of the total landings (2003: $63 \%$, 2004: $52 \%$, 2005: 57 \% and 2006: $64 \%$ ). The change in fishing pattern was caused by requirements for a fish factory on Rügen Island established in 2003 which can process 50000 t per year.

Investigation of new Danish fleet/metier description and the possibilities of improving the advice for the mixed stocks in IIIa (The IMHERSKA EU-project (Clausen et al., 2006))

An ecosystem approach to fisheries management should consider conservation of intraspecific variation due to population structure and life history variation. Knowledge of stock integrity is of unequivocal importance for sustainable fisheries management, since variable compositions in mixed areas together with asynchronous population dynamics may lead to over-fishing of individual stocks if not all components are managed to ensure (or achieve) sustainable exploitation.

A descriptive analysis of the Danish fleet dynamics during the last decade, in terms of the distribution of herring catches over fleets and at the overall activity of the vessels targeting herring in Division IIIa, together with an investigation of the fleet/metier specific exploitation of the individual stocks in Division IIIa was performed in the IMHERSKA EU project (Clausen et al., 2006).

For the descriptive analysis of the Danish fleet dynamics during the last decade, the fisheries identified in Ulrich and Andersen (2004) was modified accordingly, to get as much consistency with the previous HAWG work. Fisheries were identified using a 3-steps method using multivariate analysis of landings profile (target species) and trips descriptors (mesh size, season, and area). The data were based on logbook data and though considerable misreporting is suspected to take place between Division IIIa and the North Sea, the geographical patterns described below is believed to illustrate the fishery behaviour in general terms.

Figure A.2.1 illustrates the distribution of Danish herring landings in Division IIIa by vessel type and homeport (fleet) in 2004. From this 4 fleets were identified and Figure 3.1.2 shows the distribution of herring landings by fleet over selected years:
(1) OTB_NSSK: trawlers from North Sea and Skagerrak harbours (Skagen included). This fleet is referred to as the Northern fleet.
(2) PSB_NSSK: purse-seines from North Sea and Skagerrak harbours.
(3) OTB_KAWB: trawlers from North Sjælland and Western Baltic (Subdivisions 2224) harbours. This fleet is referred to as the Southern fleet.
(4) OTH: all other vessels recorded for having caught herring in Division IIIa at least once a year. Given its low importance, this fleet is not kept further in the analysis.


Figure A.2.1 Danish landings in IIIa by vessel and homeport.
The spatial and temporal distribution of the two main stocks (NSAS and WBSS respectively) in the SubDivisions IVaE, IIIaN, IIIaS and SubDivisions 22-24 based on analysis of herring catch compositions from both commercial and scientific sampling in the period from 1999 to 2004 appear to be following certain patterns in terms of seasonality which in turn allow predictions of the mix of herring in the area. Furthermore, by using the above four fleets/metiers and disaggregating those further into industrial or commercial activities and looking at the stock composition in their catches within different seasons, stock selective metiers was identified (a stock selective metier was defined as: a metier with $80 \%$ or more of its landings constituting the same stock). Identifying such patterns, both in terms of the lifestage spatiality of WBSS and NSAS in division IIIa and adjacent areas, and in terms of fleets activity and inter-stock selectivity was a necessary prerequisite for any use of improved fleetand stock-based management objectives. We have thus demonstrated that a more precise advice for the mixed stock in IIIa using elaborate fleet- and stock-based disaggregation could be implemented. A projection method for predicting both stock- and metier-specific Fs is being developed accordingly.

## Historical Danish fishing pattern

The general dynamics of the Danish herring activities in Division IIIa can be summed up as the following points:

- During the first half of the 1990-ties, the activity was relatively local. The fleets were mostly fishing in their immediate waters. For some of the vessels mainly participating in the small meshed fisheries the fishery for herring for human consumption was a minor but stable activity.
- The second half of the 1990-ties was a period of extension. Both the Southern and Northern trawling fleets extended their activity to the Baltic, and decreased meanwhile their industrial activities in the Kattegat and Skagerrak, which induced reduced by-catches of herring. In the same period, the large purse seiners (most of the vessels are polyvalent) increased significantly their geographical mobility, with a majority of their effort being spent outside the traditional Danish fishing grounds in the North Sea and Division IIIa as they participated in fishery for blue whiting and Norwegian spring spawning herring.


## A.3. Ecosystem aspects

Recent results from the HERGEN research-project on herring (HERGEN, EU project QLRT 200-01370, final report) reveals an increase in genetic distance between herring populations in the Baltic and successive populations in subdivisions 24, 22, 21, and 20 and finally the North Sea where genetic distance reach a maximum constant difference to the Baltic. Further, genetic differences are larger among populations within the Division IIIa and Western Baltic than among populations in the North Sea. The results also suggests that the herring spawning in spring on local spawning areas in the fjords of both the Western Baltic, the Kattegat, and the Skagerrak should be regarded as distinct spawning populations (or sub-populations) rather than as "strayers" from the Rügen-herring population. Furthermore, the contribution of these local spring spawning populations are considerable (Bekkevold et al., 2005; HERGEN, EU project QLRT 200-01370, final report).

## B. Data

## B.1. Commercial catch

Misreporting to fishing area still occurs. There is uncertainty about where the Danish landings for human consumption, reported from Division IIIa were actually taken. There is a high probability that these catches have been taken in the North Sea. Therefore, some of these catches have been transferred to the North Sea. Lastly, some landings reported as taken in the Triangle (Gilleleje, DK - Kullen, S - Helsingborg, S - Helsingør, DK), may have been taken outside this area and listed under the Kattegat.

There is at present no information about the relevance of local herring stocks/populations in relation to the fisheries and their possible influence on the stock assessment. Recent evidence from genetic differentiation among spawning aggregations in the Skagerrak suggests a potential high representation of these local spawning stocks (Bekkevold et al., 2005). Other results suggest that at least the mature proportion of the different stock components to a large extent shares migration patterns and feeding areas (Ruzzante et al., 2006; van Deurs and Ramkaer, 2006).

## B.2. Biological parameters for the assesment

Mean weights-at-age in the catch in the $1^{\text {st }}$ quarter were used as stock weights.
The proportions of F and M before spawning was assumed constant between years. F-prop was set to be 0.1 and M-prop 0.25 for all age groups.

Natural mortality was assumed constant at 0.2 for all years and $2+$ ringers. A predation mortality of 0.1 and 0.2 was added to the 0 and 1 ringers, which resulted in an increase in their natural mortality to 0.3 and 0.5 , respectively (Table 3.6.4). The estimates of predation mortality were derived as a mean for the years 1977-1995 from the Baltic MSVPA (ICES 1997/J:2).

The maturity ogive was assumed constant between years:

| W-RINGS | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}^{+}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Maturity | 0.00 | 0.00 | 0.20 | 0.75 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 |

## B.3. Surveys

The summer Danish acoustic survey in Division IIIa is part of an annual survey covering the North Sea and Division IIIa in July-August. R/V DANA conducted the survey in Division IIIa. For each sub area the mean back scattering cross section was estimated for herring, sprat, gadoids and mackerel by the TS relationships given in the Manual for Herring Acoustic Surveys in ICES Division III, IV, and IVa (ICES 2002/G:02). Used in the final assessment.

The first joint acoustic survey was carried out with R/V 'Solea’ in Subdivisions 22-24 in October 1987. Since 1989 the survey was repeated every year as a part of an international hydracoustic survey in the Baltic. Used in the final assessment.

The IBTS $3^{\text {rd }}$ quarter survey in Div. IIIa, which is a part of the North Sea and Div. IIIa bottom trawl survey that is carried out in the $1^{\text {st }}$ and $3^{\text {rd }}$ quarter. The IBTS has been conducted annually in the $1^{\text {st }}$ quarter since 1977 and $3^{\text {rd }}$ quarters from 1991. From 1983 and onwards the survey was standardised according to the IBTS manual (ICES 2002/D:03). During the HAWG 2002 the IBTS survey data (both quarter) were revised from 1991 to 2002. Historical catch rates are heavily skewed and therefore the survey indices by winter rings $1-5$ were calculated as geometric means from observed abundances $\left(n \cdot h^{-1}\right)$ at age at trawl stations. Used in the final assessment.

The German herring larvae monitoring started in 1977 and takes place every year from March/April to June in the main spawning grounds of the spring spawning herring in the Western Baltic. These are the Greifswalder Bodden and adjacent waters. For the calculation of the number of larvae per station and area unit, the methods of Smith and Richardson (1977) and Klenz (1993) were used and projected to length-classes. Further details concerning the surveys and the treatment of the samples are given in Brielmann (1989), Müller and Klenz (1994) and Klenz (2002). No data available for 2006. Not used in the final assessment.

## B.4. Commercial CPUE

## B.5. Other relevant data

## C. Historical Stock Development

Model used: ICA
Software used: ICA Vs 1.4
Model Options chosen:
No of years for separable constraint: 5
Reference age for separable constraint: 4
Constant selection pattern model : yes
S to be fixed on last age: 1.0
First age for calculation of reference F: 3
Last age for calculation of reference F: 6
Relative weights-at-age: 0.1 for 0 -group, all others 1

Relative weights by year: all 1
Catchability model used: for all indices linear
Survey weighting: Manual all 1
Estimates of the extent to which errors in the age-structured indices are correlated across ages: all 1
No shrinkage applied
Input data types and characteristics:

| Type | Name | Year range | Age range | Variable from YEAR TO YEAR Yes/No |
| :---: | :---: | :---: | :---: | :---: |
| Caton | Catch in tonnes | 1991- last data year | 0-8+ | Yes |
| Canum | Catch-at-age in numbers | 1991- last data year | 0-8+ | Yes |
| Weca | Weight-at-age in the commercial catch | 1991- last data year | 0-8+ | Yes |
| West | Weight-at-age of the spawning stock at spawning time. | 1991- last data year | 0-8+ | Yes, assumed as the Mw in the catch first quarter |
| Mprop | Proportion of natural mortality before spawning | 1991- last data year | 0-8+ | No, set to 0.25 for all ages in all years |
| Fprop | Proportion of fishing mortality before spawning | 1991- last data year | 0-8+ | No, set to 0.1 for all ages in all years |
| Matprop | Proportion mature at age | 1991- last data year | 0-8+ | No, constant for all years |
| Natmor | Natural mortality | 1991- last data year | 0-8+ | No, constant for all years |

Presently used Tuning data:

| Type | Name | Year RaNGE | AGE RANGE |
| :--- | :--- | :--- | :--- |
| Tuning fleet 1 | Danish Acoustic Survey <br> Div. IIIa | 1989 - last year data | $2-8+$ |
| Tuning fleet 2 | German Acoustic Survey <br> SDs 22-24 | 1989 - last year data | $0-5$ |
| Tuning fleet 3 | IBTS Quarter 3 | 1991 - last years data | $1-5$ |
| $\ldots$. |  |  |  |

## D. Short-Term Projection

Model used: Age structured
Software used: MFDP Vs 1a
Initial stock size: ICA estimates of population numbers were used except for

- the numbers of 0-ringers in the last two years and the start year of the projection, where a geometric mean of the recruitment over the period of 5 years was taken
- the numbers of 1-ringers in the start of the projection, where the geometric mean over the period of 5 years excluding the last year was used

Maturity: The same values as in the assessment is used for all years
F and $M$ before spawning: The same ogive as in the assessment is used for all years
Weight-at-age in the stock: Average weight of the three last years

Weight-at-age in the catch: Average weight of the three last years
Exploitation pattern: Average weight of the three last years
Intermediate year assumptions: Status quo fishing mortality
Stock recruitment model used: None
Procedures used for splitting projected catches: Not relevant

## E. Medium-Term Projections

Model used: none
Software used

Initial stock size:
Natural mortality:
Maturity:
F and M before spawning:
Weight-at-age in the stock:
Weight-at-age in the catch:
Exploitation pattern:
Intermediate year assumptions:
Stock recruitment model used:
Uncertainty models used: none

1) Initial stock size:

2 ) Natural mortality:
3 ) Maturity:
4 ) F and $M$ before spawning:
5 ) Weight-at-age in the stock:
6 ) Weight-at-age in the catch:
7 ) Exploitation pattern:
8 ) Intermediate year assumptions:
9 ) Stock recruitment model used:

## F. Long-Term Projections

Model used: none
Software used
Maturity:
F and M before spawning:
Weight-at-age in the stock:
Weight-at-age in the catch:
Exploitation pattern:
Procedures used for splitting projected catches:

## G. Biological Reference Points

Reference points have neither been defined nor proposed for this stock.

## Risk assessment performed in 2007

To address the issue of risk assessment with respect to simulation based optimizations carried out for IIIa herring in section 3.8 we implemented the following risk definition as given in the SGRAMA report of 2006 (ICES 2006/RMC:04) which is risk in a juridical sense:

$$
\begin{align*}
\text { Risk } & =\mathrm{P}(\text { harmful event }) \times \text { severity of harmful event } \\
& =\mathrm{P}(\text { lower SSB limit undercut }) \times \mathrm{EL} \tag{1}
\end{align*}
$$

with expected loss (EL) being defined as

$$
\begin{equation*}
\mathrm{EL}=\mathrm{E}\left[\mathrm{SSB}_{\text {lower limit }}-\mathrm{SSB}_{\text {estimated }} \mid \mathrm{SSB}_{\text {estimated }}<\mathrm{SSB}_{\text {lower limit }}\right] \tag{2}
\end{equation*}
$$

While this definition of risk is not only implemented as part of many national constitutions (for instance, of the German constitution; Schuldt 1997, Schulte 1999, Schulz et al. 2001) but is also commonly used in engineering, in natural or environmental sciences or in medicine (see, for instance, Burgmann 2004), in mathematical sciences however $P$ (harmful event) is often solely used as a definition for risk. As we aim at specifying costs or loss from a political and economic perspective, Eq. (1) turns out to be the appropriate risk measure, as it contains a probability term specifying the chance or likelihood of a harmful event and a severity term quantifying the magnitude of the loss. Further information on the theory underlying risk assessment and risk management can be found in Burgmann (2004), Francis and Shotton (1997) and Lane and Stephenson (1997). For a formal treatment of quantitative risk assessment and management see McNeil (2005).

## H. Other Issues

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## Quality Handbook ANNEX: Herring VIaS and VIIbc

Stock specific documentation of standard assessment procedures used by ICES.

## Stock:

Working Group:

Herring in VIaS and VIIbc
Herring Assessment Working Group for the area south of 62 N .

## Date:

## A. General

## A.1. Stock Definition

The herring to the northwest of Ireland comprise both autumn and winter spawning components. For the purpose of stock assessment and management, these areas have been separated from VIaN since 1982. Spawning in VIIb has traditionally taken place in the autumn and in VIaS, later in the autumn and in the winter.

## A.2. Fishery

The TAC is taken mainly by Ireland, which has over $90 \%$ of the quota. In recent years, only Ireland has exploited herring in this area. In 2000 the Irish North West Pelagic Management Committee was established to deal with the management of this stock. Landings have decreased markedly from about $44,000 \mathrm{t}$ in 1990 to around $13,000 \mathrm{t}$ in 2004. Landings have increased in 2006 in line with increases in the TAC.

## B. Data

## B.1. Commercial Catch

The commercial catches are provided by national laboratories belonging toe the nations that have quota for this stock. In recent years, only Ireland has caught herring in this area, so catch-at-age, mean weights and stock weights are derived entirely from Irish sampling. Sampling is performed as part of commitments under the EU Council Regulation 1639/2001.

Commercial catch at age data are submitted in Exchange sheet v 1.6.4. These data are usually processed using SALLOCL. This program (Patterson, 1998) gives outputs on sampling status and available biological parameters and documents actions taken to raise unsampled metiers using other data sets. The species co-ordinator allocates samples of catch numbers, mean length and mean weight-at-age to unsampled catches using appropriate samples by gear (fleet) area quarter and if an exact match is not available then a neighbouring area if the fishery extends to this area in the same quarter.

## B.2. Biological

Landings data are available for this area from 1970. Data on catch numbers at age, mean weights at age and mean lengths at age are derived from Irish data. The data are obtained from market sampling and are processed as described above.

## B.3. Survey

## Acoustic Surveys

Acoustic surveys have been carried out in this area since 1994. The timing of these surveys has changed over this time. Initially the surveys were undertaken in the summer in order to coincide with international herring surveys and with the summer feeding period of this stock. In 1997, a research vessel was not available and the survey was not carried out. From 1998 2001 surveys were undertaken in October in order to survey the autumn spawning component. This was changed in 2002 with surveys carried out in January targeting the winter spawning components of this stock. The acoustic survey was first used to tune the assessment in 2006.

## Larval Surveys

Assessment of this stock was largely based on the results of larval surveys in the 1980s. Herring Larval surveys were first carried out on this stock by Ireland in 1981 and continued until 1986. Prior to this the surveys were carried out by the Scottish but only had limited coverage of the assessment area. The survey grid consisted of sampling stations about 18 km apart. A gulf III plankton sampler with $275 \mu \mathrm{~m}$ mesh was towed at each station. The samples collected were preserved in $4 \%$ formalin. Herring larvae were identified and measured. Only larvae of less than 10 mm are used for the assessment. The number of larvae below each square metre was calculated and then multiplied by the area of the sea at each station (Grainger and McArdle, 1985). These surveys did not produce an index of stock size but they did provide valuable information on the distribution of very small larvae and on the location of the spawning grounds (Anon, 2000).

## Ground Fish Survey

The IGFS is part of the western IBTS survey and has been carried out on the RV Celtic Explorer since 2003. The gear used on the survey is a GOV 36/47 demersal trawl with a 20 mm cod end liner to retain juvenile and small fish, including small herring. This survey has been conducted since the early 1990s but is of little utility as a herring recruit index, because the gear, timing and survey vessel changed throughout. Once a sufficient time series becomes available it will be investigated as a possible tuning fleet. The Scottish groundfish survey, which has some coverage of VIaS will also be investigated as an additional tuning fleet.

## Herring Tagging

A herring tagging experiment was carried out in 1992 in order to investigate the movements and annual migrations of herring around the Irish Coast. 20,000 herring were tagged in total with 10,000 of these off the west coast. Some fish moved northwards and were recaptured along the north coast between July and February, in the main fishing areas. $90 \%$ of the fish tagged along the west coast were recovered from the Donegal Bay area. The maturity stages of the recaptured fish, suggests that the fish were migrating inshore towards spawning grounds (Molloy, et al 1993).

## C. Historical Stock Development

## Models used

In recent years the model used for this stock was a separable VPA. This was used to screen over three terminal fishing mortalities, $0.2,0.4$ and 0.6 . This was achieved using the Lowestoft VPA software (Darby and Flatman, 1994). Reference age for calculation of fishing mortality was 3-6 and terminal selection was fixed at 1, relative to age 4 (winter rings). In 2006 ICA was used for the first time and the acoustic surveys used as a tuning fleet. The results of these exploratory assessments are presented in the WG report. No final assessment has been accepted by the working group in recent years.

Input data types and characteristics:

| TYPE | NAME | YEAR <br> RANGE | AGE <br> RANGE | VARIABLE FROM YEAR <br> TO YEAR <br> YES/NO |
| :--- | :--- | :--- | :--- | :--- |
| Caton | Catch in tonnes | $1970-2006$ | $1-9$ | Yes |
| Canum | Catch at age in numbers | $1970-2006$ | $1-9$ | Yes |
| Weca | Weight at age in the commercial catch | $1970-2006$ | $1-9$ | Yes |
| West | Weight at age of the spawning stock at <br> spawning time. | $1970-2006$ | $1-9$ | Yes |
| Mprop | Proportion of natural mortality before <br> spawning | $1970-2006$ | $1-9$ | No |
| Fprop | Proportion of fishing mortality before <br> spawning | $1970-2006$ | $1-9$ | No |
| Matprop | Proportion mature at age | $1970-2006$ | $1-9$ | No |
| Natmor | Natural mortality | $1970-2006$ | $1-9$ | No |

Tuning data:

| TYPE | NAME | YEAR RANGE | AGE RANGE |
| :--- | :--- | :--- | :--- |
| Tuning fleet 1 | NWHAS | $1999-2007$ | $3-4$ |

## D. Short-Term Projection

Model used: Multi fleet Deterministic Projection (Smith, 2000).
Software used: MFDP Software

## E. Medium-Term Projections

Model Used: Multi Fleet Yield Per Recruit
Software Used: MFYPR Software
Yield-per-recruit analysis was carried out using MFYPR to provide yield-per-recruit plots for the data produced in the assessment. The values for $\mathbf{F}_{0.1}$ and $\mathbf{F}_{\text {med }}$ are 0.17 and 0.31 . $\mathrm{F}_{\max }$ is undefined and this is consistent with many other pelagic species.

## F. Long-Term Projections

Not performed

## G. Biological Reference Points

$\mathrm{B}_{\mathrm{pa}}$ is set at $81,000 \mathrm{t}$ and $\mathrm{B}_{\text {lim }}$ at $110,000 \mathrm{t} . \mathrm{F}_{\mathrm{pa}}$ is a 0.22 and $\mathrm{F}_{\text {lim }}$ at 0.33 .

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## Quality Handbook

ANNEX: hawg-her47d3
Stock specific documentation of standard assessment procedures used by ICES.

Stock: $\quad$ North Sea Autumn Spawning Herring (NSAS)<br>Working Group: Herring Assessment WG for the Area south of $62^{\circ} \mathrm{N}$<br>Date: 17 March 2007<br>Authors: C. Zimmermann (ed.), J. Dalskov, M. DickeyCollas, H. Mosegaard, P. Munk, J. Nichols, M. Pastoors, N. Rohlf, E.J. Simmonds, D. Skagen

## A. General

A.1. Stock definition: Autumn spawning herring distributed in ICES area IV, Division IIIa and VIId. Mixing with other stocks occurs especially in Division IIIa (with Western Baltic Spring Spawning herring).

## A.2. Fishery

North Sea Autumn Spawners are exploited by a variety of fleets, ranging from small purse seiners to large freezer trawlers, of different nations (Norway, Denmark, Sweden, Germany, The Netherlands, Belgium, France, UK, Faroe Islands). The majority of the fishery takes place in the Shetland-Orkney area in the $2^{\text {nd }}$ and $3^{\text {rd }}$ quarter, and in the English Channel (Division VIId) in the $4^{\text {th }}$ quarter. Juveniles are caught in Division IIIa and as by-catch in the industrial fishery in the central North Sea. For management purposes, 4 fleets are currently defined: Fleet A is harvesting herring for human consumption in IV and VIId, but includes herring bycatches in the Norwegian industrial fishery; fleet B is the industrial (small mesh, $<32 \mathrm{~mm}$ mesh size) fleet of EU nations operating in IV and VIId. North Sea Autumn spawners are also caught in IIIa in fleets C (human consumption) and D (small mesh).

## A.3. Ecosystem aspects:

Herring is the key pelagic species in the North Sea and is thus considered to have major impact as prey and predator to most other fish stocks in that area.

The North Sea is semi-enclosed and situated on the continental shelf of North-western Europe and is bounded by England, Scotland, Norway, Sweden, Denmark, Germany, the Netherlands, Belgium and France. It covers an area of $745,950 \mathrm{~km}^{2}$ of which the greater part is shallower than 200 m . It is one of the most diverse coastal regions in the world, with a variety of coastal habitats (fjords, estuaries, deltas, banks, beaches, sandbanks and mudflats, marshes, rocks and islands), and four ecological seasons. It is a highly productive ( $>300 \mathrm{gC} \mathrm{m}^{-2} \mathrm{yr}^{-1}$ ) ecosystem but with primary productivity varying considerably across the sea. The highest values of primary productivity occur in the coastal regions, influenced by terrestrial inputs of nutrients, and in areas such as the Dogger Bank and tidal fronts. Changes observed in trophic structure are indicative of a trend towards a decreasing resilience of this ecosystem. This trend is partially a response to inter-annual changes in the physical oceanography of the North Atlantic.

Herring are an integral and important part of the pelagic ecosystem in the North Sea. As plankton feeders they form an important part of the food chain up to the higher trophic levels.

Both as juveniles and as adults they are an important source of food for some demersal fish and for sea mammals. Over the past century the top predator, man, has exerted the greatest influence on the abundance and distribution of herring in the North Sea. Spawning stock biomass has fluctuated from estimated highs of around 4.5 million tonnes in the late 1940s to a lows of less than 100,000 tonnes in the late 1970s. The species has demonstrated a robustness in relation to recovery from such low levels once fishing mortality is curtailed in spite of recruitment levels being adversely affected.

Their spawning and nursery areas, being near the coasts, are particularly sensitive and vulnerable to anthropogenic influences. The most serious of these is the ever increasing pressure for marine sand and gravel extraction. This has the potential to seriously damage and destroy the spawning habitat and disturb spawning shoals and destroy spawn if carried out during the spawning season. Similarly, trawling at or close to the bottom in known spawning areas can have the same detrimental effects. It is possible that the disappearance of spawning on the western edge of the Dogger bank could well be attributable to such anthropogenic influences.

In more recent years the oil and gas exploration in the North Sea has represented a potential threat to herring spawning although great care has been taken by the industry to restrict their activities in areas and at times of known herring spawning activity.

## By-catch and Discard

By-catch consists of the retained 'incidental' catch of non-target species and discard is a deliberately (or accidentally) abandoned part of the catch returned to the sea as a result of economic, legal, or personal considerations. This section therefore deals with these two elements of the fishery, looking specifically at fishery-related issues. Cetacean, seabird and other threatened, rare and iconic species which may form part of a by-catch are considered separately in the next section.

Incidental Catch: The incidental catch of non-target species in the North Sea pelagic herring fishery in general is considered to be low. A recent study (Pierce et al, 2002) investigated incidental catch from commercial pelagic trawlers over the period January to August 2001. The target species, herring, accounted for $98 \%$ by weight of the overall catch with an overall incidental catch of $2.3 \%$. Mackerel, which are known to occur in mixed schools with herring in division IVa was the main by-catch species, accounting for $69 \%$ of by-catch by weight. Haddock ( $25.7 \%$ of by-catch by weight), horse mackerel Trachurus trachurus (4.8\%) and whiting Merlangius merlangus ( $0.4 \%$ ) were all present in samples. However, onboard sampling over 2002 by Scottish and German observers found substantial discards of herring, taken as by-catch in the mackerel fishery over the 3 rd and 4th quarters, after herring quotas had been exhausted.

Discards and slipping: The indications are that large-scale discarding is not widespread in the directed North Sea herring fishery. A number of direct-observer surveys have recently been conducted on Scottish and Norwegian pelagic trawlers, based on observation of 222 hauls catching 9,889 tonnes fish (Napier et al, 2002) over 2000-2002. The overall discard rate was $4.2 \%$, although that from pelagic trawlers of $6.6 \%$ was substantially higher than that from pursers $(0.6 \%)$. These discard rates were higher than the overall figure of $2.8 \%$ recorded in an earlier study (Napier et al, 1999) which were evenly distributed between pursers and trawlers. Recent surveys of Dutch vessels show a discard rate of $5 \%$ of the catch. This indicates that the different discard rates between the different fishing types in the later study were more a function of fishing location and stock size compositions rather than any gear-specific size selectivity. Some discarding, in the form of wastage (i.e. fish left meshed in the net or in the cod-end of trawls), was associated with almost all pelagic catches but the actual quantities of fish involved were low ( $2 \%$ of total discarded fish). In both studies by Napier et al., most of
the observed discarding occurred through slipping, i.e. opening the net and releasing the fish before they were pumped on-board. This occurred when catch volumes were too small, or the size of fish was too small or the fish were poor in quality. For both pursers and trawlers 'poor' fish quality was a significant cause of discarding. The size of the catch was also a significant cause of discarding from trawlers, either because the catch was too small or too large, with boats either discarding a small proportion or all of the catch. The recent influence of strong herring year classes was apparent in the composition of discards with smaller, younger fish accounting for a high proportion of the fish discarded in 2001. However surveys on the reasons why vessels discarded fish showed that larger discarding events (i.e. those $>500 \mathrm{~kg}$ ) were equally likely to the fish being of poor quality (trawlers) or the catch exceeded the vessel's capacity or market requirements (pursers). No data on survival of discarded fish has been collected but it is considered likely that mortality rates will be significant.

Ecosystem Considerations. The incidental non-target fish catch by directed North Sea herring fisheries appears to be low (ca. 2\%), mainly consisting of mackerel when fishing mixed shoals. This infers that the ecosystem level implications of incidental fish catches are negligible. The discard of unwanted herring, mostly in the form of high-grading to improve catch quality and grade sizes of fish between 2-4 years of age (see Section above) is also low, being around 3,250 tonnes (2000) and 750 tonnes (2001) for the Scottish and Norwegian and Scottish pursers and refrigerated seawater tank (RSW) pelagic trawlers operating in ICES division IVa. For both years, this was equivalent to about $10.4 \%$ by weight of the total landings. Of more concern are discards of herring from other pelagic fisheries, especially that for mackerel, where more substantial discarding of herring occurs when quotas for herring are exhausted. National reports to ICES over 1996 to 2002 suggest that total herring discards have varied between 1,500 tonnes to an unprecedented 17,000 tonnes in 2002 (reflecting onboard sampling by Scotland and Germany that observed substantial discards of herring in the mackerel fishery in the 3rd and 4th quarter in Division IVa (W)). Assuming a distribution and yield of the international mackerel fishery in IVa in 2002 to be similar to that in 2001, herring discards of all fleets could be as high as $50,000 \mathrm{t}$. This would increase the total catch in the North Sea by almost $15 \%$ and would certainly have an influence on the North Sea autumn spawning stock assessment and the perception of stock size. Discarding behaviour appears to have changed again in 2003, when herring TAC has been increased by $50 \%$, and at the same time the mackerel TAC has been reduced by more than $5 \%$.

Interactions with Rare, Protected or Icon Species: Interactions between the directed North Sea herring fishery with rare, protected or icon species are, in general, considered to be exceptional. Species which may interact with the fishery are considered below.

Cetacean by-catch: Since 2000, the Sea Mammal Research Unit (SMRU) of St. Andrew's University in Scotland, under contract to DEFRA, has carried out a number of surveys to estimate the level of by-catch in UK pelagic fisheries. SMRU, in collaboration with the Scottish Pelagic Fishermen's Association, placed observers on board thirteen UK vessels for a total of 190 days at sea, covering 206 trawling operations around the UK. To date, no cetacean by-catch has been observed in the herring pelagic fishery in the North Sea. There is currently an ongoing observer programme in the UK monitoring cetacean by-catch rates in pelagic trawl fisheries with results due at the end of September 2003 and it is understood that this confirms that cetacean by-catch by the pelagic trawl fishery is negligible (Northridge, pers. comm.). Pierce (2002) also reports that no by-catches of marine mammals were observed over 69 studies hauls and considers that the underlying rate for marine mammals in the pelagic fisheries studies (pelagic trawls in IVa and VIa) is no more than 0.05 (i.e. five events per 100 hauls) and may well be considerably lower than this.

Other than the above, there are no reliable estimates of by-catch for pelagic trawl fisheries, though observations have been made and by-catch rates have been established for several fisheries. Kuklik and Skóra (2003) refer to a single record of a harbour porpoise (Phocoena
phocoena) bycaught in a herring trawl in the Baltic. Observations in several other pelagic trawl fisheries were reported by Morizur et al. (1999) and Couperus (1997). All appear to agree that incidental catches of cetaceans in the Dutch pelagic trawl fishery are largely restricted to late-winter/early-spring in an area along the continental slope southwest of Ireland.

On 24 July 2003 the European Commission issued a proposal for a Council Regulation to address the problem of cetacean by-catch in various fisheries. For the North Sea (ICES IV) $5 \%$ of pelagic trawl fisheries would have to be monitored by observers. In the eastern channel $5 \%$ of pelagic trips would have to be monitored from April to November but $10 \%$ from December to March. The Commission has asked the Council to adopt this proposal by 1 July 2004.

Seal by-catch: The by-catch of seals in directed pelagic herring fishery in the North Sea is reported to be "very rare" (Aad Jonker, pers. comm.). Independent verification also confirms this to be so, with perhaps one animal being caught by the whole North Sea fleet a year (Bram Couperus (RIVO), pers. comm.). Northridge (2003) observed 49 seals taken in 312 pelagic trawl tows throughout UK waters and reports that the fishery in North-western Scotland has the highest observed seal by-catch levels of UK pelagic trawl fisheries, possible amounting to dozens per year. Although not confirmed, it was assumed that the majority were grey seal Halichoerus grypus. This species is mainly distributed around the Orkneys and Outer Hebrides - out of a UK population of 129,000, only around 7,000 and 5,900 are distributed off the Scottish and English North Sea coasts respectively (SCOS, 2002), and so by-catch rates in the North Sea are likely to be substantially less than off the NW Scottish coast. The eastern Atlantic population of the Grey seal is not considered to be threatened.

Other by-catch: Sharks are occasionally caught by pelagic trawlers in the North Sea, although this is rare with a maximum of two fish per trip (Aad Jonker, pers. comm.). Survival rates are apparently high, with sharks being released during or after the cod-end is being emptied. The species are unknown, although blue shark Prionace glauca, which preys primarily upon schooling fishes such as anchovies, sardines, herring, are known to have been caught by pelagic trawls off the SW English coast (Bram Couperus (RIVO), pers. comm.). Gannets (Morus bassanus), which frequently dive at and around nets, were observed by Napier et al. (2002) entangled in the nets but were not present in samples. Actual mortality rates of caught gannets have not been assessed in detail, and some have been observed alive after release from the gear. An extrapolation from observed mortalities corresponds to around 560 gannet deaths per year, although this is based on a relatively low sample frame. Seabird by-catch in the North Sea is considered to be comparatively rare compared to the NW Scotland where 1-3 birds may be caught, esp. in grounds off St. Kilda (Aad Jonker (former freezer trawler skipper), pers. comm.). RIVO observers in the North Sea only recorded one incident of seabird by-catch over 10 trips (Bram Couperus (RIVO), pers. comm.).

## B. Data

## B.1. Commercial catch:

Commercial catch is obtained from national laboratories of nations exploiting herring in the North Sea. Since 1999 (catch data 1998), these labs have used a spreadsheet to provide all necessary landing and sampling data, which was developed originally for the Mackerel Working Group (WGMHSA) and further adapted to the special needs of the Herring Assessment Working Group. The current version used for reporting the 2003 catch data was v1.6.4. The majority of commercial catch data of multinational fleets was provided on these spreadsheets and further processed with the SALLOCL-application (Patterson, 1998). This program gives the needed standard outputs on sampling status and biological parameters. It also clearly documents any decisions made by the species co-ordinators for filling in missing
data and raising the catch information of one nation/quarter/area with information from another data set.

Transparency of data handling by the Working Group. The current practice of data handling by the Working Group is that the data received by the co-ordinators is available in a folder called "archive". These high-resolution data are not reproduced in the report. The archived data contains the disaggregated dataset (disfad), the allocations of samples to unsampled catches (alloc), the aggregated dataset (sam.out) and (in some cases) a document describing any problems with the data in that year.

Current methods of compiling fisheries assessment data. The species co-ordinator is responsible for compiling the national data to produce the input data for the assessments. In addition to checking the major task involved is to allocate samples of catch numbers, mean length and mean weight-at-age to unsampled catches. There are at present no defined criteria on how this should be done, but the following general process is implemented by the species co-ordinators. Searches are made for appropriate samples by gear (fleet) area quarter, if an exact match is not available the search will move to a neighbouring area if the fishery extends to this area in the same quarter. More than one sample may be allocated to an unsampled catch, in this case a straight mean or weighted mean of the observations may be used. If there are no samples available the search will move to the closest non-adjacent area by gear (fleet) and quarter, but not in all cases.

The Working Group acknowledges the effort some members have made to provide "corrected" data, which in some cases differ significantly from the officially reported catches. Most of this valuable information is gathered on the basis of personal knowledge of the fishery and good relations between the scientist responsible and the fishermen. The WG is aware of the problem that this knowledge might be lost if the scientist leaves, and asks the national laboratories to ensure continuity in data provision. In addition the Working Group recognises and would like to highlight the inherent conflict of interest in obtaining details of unallocated catches by country and increasing the transparency of data handling by the Working Group. This issue will have to be carefully considered in light of any future development by ICES of a standard platform to store all fisheries disaggregated data, particularly with regard to confidentiality.

The WG considered the need of a long-term data storage for commercial catches and sampling, and the documentation of any primary data processing of these data. From 2000 on (catch data for 1999), the latest (consistency checked) versions of the input files together with standard outputs and a documentation of filling-in decisions made by the co-ordinators, ideally in the SALLOC-formats, are stored in a separate "archive" folder. This is updated annually, and the complete collection (which is supposed to be kept confidential as it will contain data on misreporting and unallocated catches) will be available for WG members on request. As there was very little historical information available, WG members were asked to provide as much as possible national catch and historical data sets in any available format which is then stored in a " $\sim$ historic" folder within "Archive". They will be consistency checked and transferred into a database system as soon as this is available. In 2007, INTERCATCH was used for the first time and compared to the SALLOC formats, no major differences were found.

## B.2. Biological

Catch-at-age data (catch numbers-at-age, mean weights-at-age in the catch, mean length-atage) is derived from the raised national figures received from the national laboratories. The data is obtained either by market sampling or by onboard observers, and processed as described above. For information on recent sampling levels and nations providing samples, see Sec. 2.2. of the most recent HAWG report.

Mean weights-at-age in the stock and proportions mature (maturity ogive) are derived from the June/July international acoustic survey (see next paragraph).

## B.3. Surveys

## B.3.1 Acoustic: ICES Co-ordinated Acoustic Surveys for herring in North Sea, Skagerrak and Kattegat

The ICES Coordinated acoustic surveys started in 1979 around Orkney and Shetland with first major coverage in 1984. An index derived from that survey has been used in assessments since 1994 with the time-series data extending back to 1989. The survey was extended to IIIa to include the overlapping Western Baltic spring spawning stock in 1989, and the index has been used with a number of other tuning indices since 1991. The early survey had occasionally covered VIa (North) during the 1980s and was extended westwards in 1991 to cover the whole of VIa (North) annually since 1991, and provides the only tuning index for VIa (North) herring, By carrying out the co-ordinated survey at the same time from the Kattegat to South of the Hebrides all herring in these areas are covered simultaneously, reducing uncertainly due to area boundaries as well as providing input indices to three distinct stocks. The surveys are co-ordinated under ICES Planning Group for Herring Surveys ICES PGHERS.

At present, six surveys are carried out during late June and July covering most of the continental shelf north of $52^{\circ} \mathrm{N}$ in the North Sea and to the west of Scotland to a northern limit of $62^{\circ} \mathrm{N}$. The eastern edge of the survey area is bounded by the Norwegian and Danish, Swedish and German coasts, and to the west by the shelf edge between 200 and 400 m depth. The surveys are reported individually in the report of the planning group for herring surveys, and a combined report is prepared from the data from all surveys. The combined survey results provide spatial distributions of herring abundance by number and biomass at age by statistical rectangle; and distributions of mean weight and fraction mature at age.

The acoustic recordings are carried out using Simrad EK60, EK500 and EY500 38 kHz sounder echo-integrator with transducers mounted on the hull, drop keel or towed bodies. Further data analysis is carried out using either BI500, Echoview or Echoann software. The survey track is selected to cover the area giving a basic sampling intensity over the whole area based on the limits of herring densities found in previous years. A transect spacing of 15 nautical miles is used in most parts of the area with the exception of some relatively high density sections, east and west of Shetland, in the Skagerrak where short additional transects were carried out at 7.5 nmi spacing, and in the southern area where a 30 nmi transect spacing is used.

The following target strength to fish length relationships have been used to analyse the data:

$$
\begin{array}{ll}
\text { herring } & \mathrm{TS}=20 \log \mathrm{~L}-71.2 \mathrm{~dB} \\
\text { sprat } & \mathrm{TS}=20 \log \mathrm{~L}-71.2 \mathrm{~dB} \\
\text { gadoids } & \mathrm{TS}=20 \log \mathrm{~L}-67.5 \mathrm{~dB} \\
\text { mackerel } & \mathrm{TS}=21.7 \log \mathrm{~L}-84.9 \mathrm{~dB}
\end{array}
$$

Data is reported through standardised data exchange format and combined at FRS Marine Lab Aberdeen. The exchange format currently holds information on the ICES statistical rectangle level, with at least one entry for each rectangle covered, but more flexible strata are accommodated by allowing multiple entries for abundance belonging to different strata. Data submitted consists of the ICES rectangle definition, biological stratum, herring abundance by proportion of Autumn spawners (North Sea and VIa North) and Spring spawners (Western Baltic, age and maturity, and survey weight (survey track length). Data are be presented according to the following age/maturity classes: 1 immature (maturity stage 1 or 2 ), 1 mature (maturity stage $3+$ ), 2 immature, 2 mature, 3 immature, 3 mature, 4, 5, 6, 7, 8, 9+. In addition to proportions at age data on mean weights and mean length are reported at age/maturity by biological strata. Data is combined using an effort weighted mean based on survey effort reported as number of nautical miles of cruise track per statistical rectangle. A combined survey report is produced annually. Apart from the Biomass index for 1-9+-ringers,mean weights at age in the catch and proportions mature are derived from the survey to be used in the NSAS assessment.

## B.3.2 International Bottom Trawl Survey:

The International Bottom Trawl Survey (IBTS) started out as a Young Herring Survey (IYHS) in 1966 with the objective of obtaining annual recruitment indices for the combined North Sea herring stocks. It has been carried out every year since, and it was realized that the survey could provide recruitment indices not only for herring, but for roundfish species as well. Examination of the catch data from the 1st quarter IBTS showed that these surveys also gave indications of the abundances of the adult stages of herring, and subsequently the catches have been used for estimating 2-5+ ringer abundances. The surveys are carried out in $1^{\text {st }}$ quarter (February) and in $3^{\text {rd }}$ quarter (August-September) using standardized procedures among all participants. The standard gear is a GOV trawl, and at least two hauls are made in each statistical rectangle.

In 1977 sampling for late stage herring larvae was introduced at the IBTS $1^{\text {st }}$ quarter, using Isaccs-Kidd Midwater trawls. These catches appeared as a good indicator of herring recruitment, however examination of IKMT performance showed deficiencies in its catchability for herring larvae, and a more applicable gear, a ring net (MIK) was suggested as an alternative gear. Hence, gear type was changed in the mid 90'ies, and the MIK has been the standard gear of the program since. This ring net is of 2 meter in diameter, has a long twolegged bridle, and is equipped with a black netting of 1.5 mm mesh size. Oblique hauls are made during night in at least two statistical rectangles.

Indices of 2-5+ ringer herring abundances in the North Sea ( $\mathbf{1}^{\text {st }}$ quarter). Fishing gear and survey practices were standardised from 1983, and herring abundance estimates of 2-5+ ringers from 1983 onwards has shown the most consistent results in assessments of these age groups. This series is used in North Sea herring assessment. The catches in DivisionIIIa is not included in this index. Table 2.3.3.1 in the HAWG report shows the time series of abundance estimates of 2-5+ ringers from the 1st quarter IBTS for the whole period.

Index of 1-ringer recruitment in the North Sea ( ${ }^{\text {st }}$ quarter). The 1-ringer index of recruitment is based on trawl catches in the entire survey area, hence, all 1-ringer herring caught in Div IIIa is included in this index. Indices are calculated as an area weighted mean over means by ICES statistical rectangle, and are available for year classes 1977 to recent (Table 2.3.3.3 of HAWG report). The Downs herring hatch later than the other autumn spawned herring and generally appears as a smaller sized group during the 1st quarter IBTS. A recruitment index of smaller sized 1-ringers is calculated using the standard procedure, but solely based on abundance estimates of herring $<13 \mathrm{~cm}$ (see discussion of procedures in earlier reports (ICES CM 2000/ ACFM:10, and ICES CM 2001/ ACFM:12).

MIK index of 0-ringer recruitment in the North Sea ( $1^{\text {st }}$ quarter). The MIK catches of late stage herring larvae is used to calculate and 0-ringer index of autumn spawned herring in the North Sea. A flowmeter at the gear opening is used for estimation of volume filtered by the gear, and using this information together with information on bottom depth, the density of herring larvae per square meter is estimated. A mean herring density in statistical rectangles is raised to mean within subareas, and based on areas of these subareas an index of total abundance is estimated (see also ICES 1996/Asses:10). The series of estimates for subareas as well as the total index are shown in the actual report's Table 2.3.3.4.

## B.3.3. Larvae:

Surveys of larval herring have a long tradition in the North Sea. Sporadic surveys started around 1880, and available scientific data goes back to the middle of the 20th century. The coordination of the International Herring Larvae Surveys in the North Sea and adjacent waters (IHLS) by ICES started in 1967, and from 1972 onwards all relevant data are achieved in a data base. The surveys are carried out annually to map larval distribution and abundance. Larval abundance estimates are of value as relative indicators of the herring spawning biomass in the assessment.

Nearly all countries surrounding the North Sea have participated in the history of the IHLS. Most effort was undertaken by the Netherlands, Germany, Scotland, England, Denmark and Norway. A number of other nations have contributed occasionally. A sharp reduction in ship time and number of participating nations occurred in the end of the 1980s. Since 1994 only the Netherlands and Germany contribute to the larvae surveys, with one exception in 2000 when also Norway participated.

Larvae Abundance Index (LAI): The total area covered by the surveys is divided into 4 sub areas corresponding to the main spawning grounds. These sub areas have to be sampled in different given time intervals. The sampling grid is standardized and stations are approximately 10 nautical miles apart. The standard gear is a GULF III sampler or one of its national modifications. Newly hatched larvae less than 10 mm total length ( 11 mm for the Southern North Sea) are used in the index calculation. To estimate larval abundance, the mean number of larvae per square meter obtained from the Ichthyoplankton hauls is raised to rectangles of $30 \times 30$ nautical miles and the corresponding surface area. These values are summed up within the given unit and provide the larval abundance per unit and time interval.

Multiplicative Larval Abundance Index (MLAI): The traditional LAI and LPE (Larval Production Estimates) rely on a complete coverage of the survey area. Due to the substantial decline in ship time and sampling effort since the end of the 80s, these indices could not be calculated in their traditional form since 1994. Instead, a multiplicative model was introduced for calculating a Multiplicative Larvae Abundance Index (MLAI, Patterson \& Beveridge, 1995). In this approach the larvae abundances are calculated for a series of sampling units. The total time series of data is used to estimate the year and sampling unit effects on the abundance values. The unit effects are used to fill unsampled units so that an abundance index can be estimated for each year.

Calculation of the linearised multiplicative model was done using the equation:

$$
\ln (\text { Indexyear,LAI unit })=\text { MLAIyear }+ \text { MLAI }_{\text {LAI }} \text { unit + uyear, LAI unit }
$$

where MLAIyear is the relative spawning stock size in each year, $M L A I_{L A I}$ unit are the relative abundances of larvae in each sampling unit and year, LAI unit are the corresponding residuals. The unit effects are converted such that the first sampling unit is used as a reference (Orkney/Shetland 01-15.09.72) and the parameters for the other sampling units are redefined as differences from this reference unit. The model is fitted to abundances of larvae less than 10
mm in length ( 11 mm for SNS). The MLAI is updated annually and represent all larval data since 1972. The time series is used as a biomass index in the herring assessment.

## B.4. Commercial CPUE

Not used for pelagic stocks.

## B.5. Other relevant data

## B.5.1 Separation of North Sea Autumn Spawners and IIIa-type Spring Spawners

North Sea Autumn Spawners and IIIa-type Spring Spawners occur in mixtures in fisheries operating in Divisions IIIa and IVaE (ICES, 1991/Assess:15): mainly 2+ ringers of the Western Baltic spring-spawners and 0-2-ringers from the North Sea autumn-spawners, including winter-spawning Downs herring. In addition, several local spawning stocks have been identified with a minor importance for the herring fisheries (ICES, 2001/ACFM 12).

The method of separating herring in Norwegian samples, using vertebral counts as described in former reports of this Working Group (ICES 1990/ Assess:14) assumes that for autumn spawners, the mean vertebral count is 56.5 and for Spring spawners 55.80. The fractions of spring spawners (fsp) are estimated from the formula (56.50-v)/(56.5-55.8), where $v$ is the mean vertebral count of the (mixed) sample with the restriction that the proportion should be one if fsp>=1 and zero if $\mathrm{fsp}<=0$. The method is quite sensitive to within-stock variation (e.g. between year classes) in mean vertebral counts.

Experience within the Herring Assessment Working Group has shown that separation procedures based on size distributions often will fail. The introduction of otolith microstructure analysis in 1996-97 (Mosegaard and Popp-Madsen, 1996) enables an accurate and precise split between three groups, autumn, winter and spring-spawners; however, different populations with similar spawning periods are not resolved with the present level of analysis. Different stock components that are not easily distinguished by their otolith microstructure (OM), are considered to have different mean vertebral counts (vs) as, e.g., winter-spawning Downs herring: 56.6 (Hulme, 1995), and the small local stocks, the Skagerrak winter/spring-spawners: 57 (Rosenberg and Palmén, 1982). Further, the estimated stock specific mean vs count varies somewhat among different studies; North Sea: 56.5, Western Baltic Sea: 55.6 (Gröger and Gröhsler, 2001) and North Sea: 56.5, Western Baltic Sea: 55.8 (ICES 1992/H:5). Comparison between separation methods using frequency distributions of vertebral counts and otolith microstructure showed reasonable correspondence. Using this information the years from 1991 to 1996 was reworked in 2001, applying common splitting keys for all years by using a combination of the vertebral count and otolith microstructure methods (ICES, 2001/ACFM:12). From 2001 and onwards, the otolith-based method only has been used for the Division IIIa.

Different methods of identifying herring stocks in the Division IIIa and Subdivisions 22-24 were recently evaluated in a EU CFP study project (EC study 98/026). The study involved several inter-calibration sessions between microstructure readers in the different laboratories involved with the WBSS herring. After the study was finished a close collaboration concerning reader interpretations has been kept between the Danish and Swedish laboratories. Sub-samples of the 2002 and 2003 Danish, Swedish, and German microstructure analyses were double-checked by the same Danish expert reader for consistency in interpretation. The overall impression is an increasingly good agreement among readers.

New molecular genetic approaches for stock separation are being developed within the EUFP5 project HERGEN (EU project QLRT 200-01370). Sampling of spawning aggregations during spring, autumn and winter has been carried out in 2002 and in 2003 in Division IIIa and in the Western Baltic at more than 10 different locations. Preliminary results point at a substantial genetic variation between North Sea and Western Baltic herring.

After the introduction of otolith microstructure analysis in 1996 it was discovered that in the western Baltic a small percentage of the herring landings might consist of autumn-spawners individuals. Before molecular genetic methods became available for Atlantic herring the existence of varying proportions of autumn spawners in Subdivisions 22-24 in different years was considered a potential problem for the assessment.

## C. Historical Stock Development

Model used:
Details on input parameters and model setup for the final ICA assessment are presented in Table 2.6.2.1. of the most recent HAWG report. The assessment has the same set-up and basic assumption as the assessment that was carried out last year. Input data are given in Tables 2.6.2.2. The ICA program operates by minimising the following general objective function:
which is the sum of the squared differences for the catches (separable model), the indices



$$
\sum_{a=1, y=1989}^{a=9+y=2002} \lambda_{a, a c o u s t}\left(\ln \left(q_{a, a c o u s t} \cdot \hat{N}_{a, y}\right)-\ln \left(\text { ACOUST }_{a, y}\right)\right)^{2}+
$$


with the following variables:

| owing variables: | es: $(\alpha \hat{S} \hat{S} R)$ |
| :---: | :---: |
| $\sum_{y=2002}^{y=1960} \lambda_{s}$ | $\left(\ln \left(\hat{N}_{\text {Pxat }}\right)-\ln \frac{\alpha S S B_{y}}{0}\right)^{2}$ |
| C | Catch at age (rings) |
| $\hat{C}$ | Estimated catch at age (rings) in the separable model |
| $\hat{N}$ | Estimated population numbers |
| $\hat{S} \hat{S} B$ | Estimated spawning stock size |
| MLAI | MLAI index (biomass index) |
| ACOUST | Acoustic index (age disaggregated) |
| IBTS | IBTS index (1-5+ ringers) |
| MIK | MIK index (0-ringers) |
| q | Catchability |
| k | power of catchability model |
| $\alpha, \beta$ | parameters to the Beverton stock-recruit model |
| $\lambda$ | Weighting factor |

Software used: ICA (Patterson, 1998; Needle, 2000)
Model Options chosen:
Input data types and characteristics:

| Type | NAME | Year RANGE | AGE <br> RANGE | VARIABLE FROM YEAR <br> TO YEAR <br> Yes/No |
| :--- | :--- | :--- | :--- | :--- |
| Caton | Catch in tonnes |  |  |  |
| Canum | Catch at age in numbers | $1960-2006$ | $1-9+$ | Yes |
| Weca | Weight at age in the commercial catch | $1960-2006$ | $1-9+$ | Yes (smoothed) |
| West | Weight at age of the spawning stock at <br> spawning time. | $1960-2006$ | $1-9+$ | Yes (smoothed) |
| Mprop | Proportion of natural mortality before <br> spawning | $1960-2006$ | $1-9+$ | No |
| Fprop | Proportion of fishing mortality before <br> spawning | $1960-2006$ | $1-9+$ | No |
| Matprop | Proportion mature at age | $1960-2006$ | $1-9+$ | Yes (smoothed) |
| Natmor | Natural mortality | $1960-2006$ | $1-9+$ | No |

Tuning data:

| Type | Name | Year range | Age range (Wr) |
| :--- | :--- | :--- | :--- |
| Tuning fleet 1 | IBTS Q1 | $1984-2007$ | $1-5$ |
| Tuning fleet 2 | MIK | $1992-2007$ | 0 |
| Tuning fleet 3 | Acoustic | $1989-2006$ | $1--9+$ |
| Tuning fleet 4 | MLAI | $1973-2006$ | SSB |
|  |  |  |  |

## D. Short-Term Projection

The short-term prediction method was substantially modified in 2002. Following the review by SGEHAP (ICES 2001/ACFM:22), which recommended that a simple multi-fleet method would be preferable, the complex split-factor method used for a number of years prior to 2002 has not been used since. The multi-fleet, multi-option, deterministic short-term prediction programme (MFSP) was accepted by ACFM and was developed further last year. It is intended to continue to use this programme in the future. The good agreement between predicted biomass for the actual year and SSB taken from the assessment for the most recent year one year after demonstrates that the current prediction procedure for stock numbers is working well. In 2004, the Working Group has included prediction of low maturation into projections for 2005 and expects to monitor growth and maturation of North Sea herring carefully in the future and when deemed necessary will include these changes in predictions in the future.

Model used: Age-structured model, by fleet and area fished
Software used: MFSP
Initial stock size: output from ICA
Maturity: average of the two most recent years used
F and M before spawning: 0.67 for both (assumes spawning starts around September)
Weight at age in the stock: from last year in assessment (already smoothed, see assessment data description)

Weight at age in the catch: average of last two years BY FLEET
Exploitation pattern:
Intermediate year assumptions: Status quo F
Stock recruitment model used: Recent average recruitment (arithmetic, recent 10 years) is used, (unless there is some strong reason for using something else, e.g. if SSB is very low, we may use a prediction from the stock-recruit relationship)

Procedures used for splitting projected catches:
There are 4 values input for this parameter:
a) IBTS 1-ringer proportion in last assessment year (y) is used for 1-ringers in y
b) IBTS 1-ringer proportion in $\mathrm{y}+1$ is used for 1-ringers in $\mathrm{y}+1$, AND for 0-ringers in $y$.
c) GLM (between MIK index and IBTS 1-ringer proportion) is applied to MIK index in $y+1$ to predict proportion for 1 -ringers in $y+2$, AND for 0 -ringers in $y+1$

GLM, as in (c), is applied to the Average MIK index for 1981 to year y to predict proportion for 1-ringers in $y+3$ (not relevant), AND for 0-ringers in $y+2$ (relevant)
E. Medium-Term Projections - still to be filled in -

Model used:
Software used:
Initial stock size:
Natural mortality:
Maturity:
F and M before spawning:
Weight at age in the stock:
Weight at age in the catch:
Exploitation pattern:
Intermediate year assumptions:
Stock recruitment model used:
Uncertainty models used:
Initial stock size:
Natural mortality:
Maturity:
F and M before spawning:
Weight at age in the stock:
Weight at age in the catch:
Exploitation pattern:
Intermediate year assumptions:
Stock recruitment model used:
F. Long-Term Projections - still to be filled in -

Model used:

Software used:
Maturity:
F and M before spawning:
Weight at age in the stock:
Weight at age in the catch:
Exploitation pattern:
Procedures used for splitting projected catches:

## G. Biological Reference Points

There is a well functioning harvest control rule in place for this stock, and apart from $\mathrm{B}_{\mathrm{lim}}$, the current reference points are derived from this HCR. The target F in the HCR was adopted by ACFM as the $\mathrm{F}_{\mathrm{pa}}$, while the trigger point at which F should be reduced below the target is adopted as $\mathrm{B}_{\mathrm{pa}}$. The HCR was briefly revisited in 2004, and the results support the initial definitions of limits.

Reference points currently in use are: $\mathrm{B}_{\mathrm{lim}}$ is 800000 t (below this value poor recruitment has been experienced); $\mathrm{B}_{\mathrm{pa}}$ be set at 1.3 mill. T (as part of a harvest control rule based on simulations); $\mathrm{F}_{\text {lim }}$ is not defined, $\mathrm{F}_{\mathrm{pa}}$ be set at $\mathrm{F}_{\text {ages } 0-1}=0.12, \mathrm{~F}_{\text {ages 2-6 }}=0.25$ (as part of a harvest control rule).

## H. Other Issues

## H. 1 Biology of the species in the distribution area

The herring (Clupea harengus) is a pelagic species which is widespread in its distribution throughout the North Sea. The herring's unique habit is that it produces benthic eggs which are attached to a gravely substrate on the seabed. This points strongly to an evolutionary history in which herring spawned in rivers and at some later date re-adapted to the marine environment. The spawning grounds in the southern North Sea are in fact located in the beds of rivers which existed in geological times and some groups of spring spawning herring still spawn in very shallow inshore waters and estuaries. Spawning typically occurs on coarse gravel $(0.5-5 \mathrm{~cm})$ to stone $(8-15 \mathrm{~cm})$ substrates and often on the crest of a ridge rather than hollows. For example, in a spawning area in the English Channel, eggs were found attached to flints $2.5-25 \mathrm{~cm}$ in length, where these occurred in gravel, over a 3.5 km by 400 m wide strip.

As a consequence of the requirement for a very specific substrate, spawning occurs in small discrete areas in the near coastal waters of the western North Sea. They extend from the Shetland Isles in the north through into the English Channel in the south. Within these specific areas actual patches of spawn can be extremely difficult to find.

The fecundity of herring is length related and varies between approximately 10,000 and 60,000 eggs per female. This is a relatively low fecundity for teleosts, probably because, in evolutionary terms, the benthic egg is a potentially less hazardous phase of development compared with the planktonic egg of most other teleosts. The age of first maturity is 3 years old (2 ringers) but the proportion mature at age may vary from year to year dependent on feeding conditions. Over the past 15 years the proportion mature at age 3 years (2 ringers) has ranged from $47 \%$ to $86 \%$ and for 4 year old fish ( 3 winter ringers) from $63 \%$ to $100 \%$. Above that age, all are considered to be mature.

The benthic eggs take about three weeks to hatch dependant on the temperature. The larvae on hatching are 6 mm to 9 mm long and are immediately planktonic. Their yolk sac lasts for a few days during which time they will begin to feed on phytoplankton and small planktonic
animals. Their planktonic development lasts around three to four months during which time they are passively subjected to the residual drift which takes them to various coastal nursery areas on both sides of the North Sea and into the Skagerrak and Kattegat.

Herring continue to be mainly planktonic feeders throughout their life history although there are numerous records of them taking small fish, such as sprat and sandeels, on an opportunistic basis. Calanoid copepods, such as Calanus, Pseudocalanus and Temora and the Euphausids, Meganyctiphanes and Thysanoessa still form the major part of their diet during the spring and summer and are responsible for the very high fat content of the fish at this time.

In the past, herring age has been determined by using the annual rings on the scales. In more recent years the growth rings on the otolith have proved more reliable for age determination. Herring age is expressed as number of winter rings on the otolith rather than age in years as for most other teleost species where a nominal 1 January birthdate is applied. Autumn spawning herring do not lay down a winter ring during their first winter and therefore remain as ' 0 ' winter ringers until the following winter. When looking at year classes, or year of hatching, it must be remembered that they were spawned in the year prior to their classification as ' 0 ' winter ringers.

North Sea herring comprise both spring and autumn spawning groups but the major fisheries are carried out on the offshore autumn spawning fish. The spring spawners are found mainly as small discrete coastal groups in areas such as The Wash and the Thames estuary. Juveniles of the spring spawning stocks found in the Baltic, Skagerrak and Kattegat may also be found in the North Sea as well as Norwegian coastal spring spawners.

The main autumn spawning begins in the northern North Sea in August and progresses steadily southwards through September and October in the central North Sea to November and as late as January in the southern North Sea and eastern English Channel. The widespread but discrete location of the herring spawning grounds throughout the western North Sea has been well known and described since the early part of the 20 th Century. This led to considerable scientific debate and eventually to investigation and research on stock identity. The controversy centred on whether or not the separate spawning grounds represented discrete stocks or 'races' within the North Sea autumn spawning herring complex. Resolution of this issue became more urgent as the need for the introduction of management measures increased during the 1950's. The International Council for the Exploration of the Sea (ICES) encouraged tagging and other racial studies and a review of all the historic evidence to resolve this problem. The conclusions were the basis for establishing the working hypothesis that the North Sea autumn spawning herring comprise a complex of three separate stocks each with separate spawning grounds, migration routes and nursery areas, illustrated in the figure below.

The three stock units are:

-     - The Buchan or Scottish group which spawn from July to early September in the Orkney Shetland area and off the Scottish east coast. Nursery areas for fish up to two years old are found along the east coast of Scotland and also across the North Sea and into the Skagerrak and Kattegat.
-     - The Banks or central North Sea group, which derive their name from their former spawning grounds around the western edge of the Dogger Bank. These spawning grounds have now all but disappeared and spawning is confined to small areas along the English east coast, from the Farne Islands to the Dowsing area, from August to October. The juveniles are found along the east coast of England, down to the Wash, and also off the west coast of Denmark.
-     - The Downs group which spawns in very late Autumn through to February in the southern Bight of the North Sea and in the eastern English Channel. The drift of their larvae takes them north-eastwards to nursery areas along the Dutch coast and into the German Bight (Burd 1985).

At certain times of the year, individuals from the three stock units may mix and are caught together as juveniles and adults but they cannot be readily separated in the commercial catches. As a consequence, North Sea autumn spawning herring have to be managed as a single unit.

A further complication is that juveniles of the North Sea stocks are found, outside the North Sea, in the Skagerrak and Kattegat areas and are caught in various fisheries there. The proportions of juveniles of North Sea origin, found in these areas varies with the strength of the year class, with higher proportions in the Skagerrak and Kattegat when the year class is good.

## H. 2 Historic stock development and history of the fishery

Over many centuries the North Sea herring fishery has been a cause of international conflict sometimes resulting in war, but in more recent times in bitter political argument. There have also been fundamental changes in the nature of the fisheries. These have been driven both by changes in catching power and in response to changes in market requirements, particularly the demand for fish meal and oil. Most of these changes have resulted in greater exploitation pressures that increasingly led to the urgent need to ensure a more rational exploitation of North Sea herring. Such pressures really began to exert themselves for the first time during the 1950's when the spawning stock biomass of North Sea autumn spawning herring fell from 5 million tonnes in 1947 to 1.4 million tonnes by 1957. That period also witnessed the decline and eventual disappearance of a traditional autumn drift net fishery in the southern North Sea.

The annual landings from 1947 through to the early 1960's were high, but stable, averaging around 650,000 t. Over the period 1952-62 the high fishing mortality (F 0.4 ages 2-6) resulted in a rapid decline in the spawning stock biomass from around 5 million tonnes to 1.5 million tonnes. Recruitment over this period was reasonable, but there were fewer and fewer year classes present in the adult stock, a clear indication that the stocks were being over-fished and that they were also being impacted by the developing industrial fishery in the eastern North Sea.

This period witnessed the complete collapse of the historic East Anglian autumn drift net fishery, which was based entirely on the Downs stock moving south to the Southern Bight and eastern English Channel to spawn. The reasons for that failure have been attributed both to high mortality of the juveniles in the North Sea industrial fisheries, and to heavy fishing by bottom trawlers on the spawning concentrations, in the English Channel, during the 1950's. Such intensive trawling, on vulnerable spawning fish, not only generated a high mortality but also disturbed spawning aggregations, destroyed the spawn and damaged the substrate on which successful spawning depends.

Fishing mortality on the herring in the central and northern North Sea began to increase rapidly in the late 1960's and had increased to F1.3 ages 2-6, or over 70\% per year of those age classes, by 1968. Landings peaked at over 1 million tonnes in 1965 , around $80 \%$ of which were juvenile fish. This was followed by a very rapid decline in the SSB and the total landings. By 1975 the SSB had fallen to 83,500t although the total landings were still over $300,000 \mathrm{t}$. At the same time, spawning in the central North Sea had contracted to the grounds off the east coast of England whilst spawning grounds around the edge of the Dogger Bank were no longer used. This heralded the serious decline and near collapse of the North Sea autumn spawning herring stock which led to the moratorium on directed herring fishing in the North Sea from 1977 to 1981.

International larvae surveys and acoustic surveys were used to monitor the state of the stocks during the moratorium. By 1980 these surveys were indicating a modest recovery in the SSB from its 1977 low point of 52,000t. By 1981 the SSB had increased to over 200,000t. Prior to the moratorium there had been no control, other than market forces, on catches in the North

Sea directed herring fishery. Once the fishery re-opened in 1981 the North Sea autumn spawning herring stock was managed by a Total Allowable Catch (TAC) constraint. It should be noted that the TAC was only applied to the directed herring fishery in the North Sea which exploited mainly adult fish for human consumption. Targeted fishing for herring for industrial purposes was banned in the North Sea in 1976 but there was a $10 \%$ by-catch allowance in the fisheries for other species, including the small meshed fisheries for industrial purposes, mainly for sprat. Following the re-opening of the now controlled fishery the SSB steadily increased, peaking at 1.3 million tonnes in 1989. Annual recruitment, measured as ' 0 'group fish, was well above the longterm average over this period. The 1985 year class was the biggest recorded since 1960 and the third highest in the records dating back to 1946. Landings also steadily increased over this period reaching a peak of 876,000 tonnes in 1988. This resulted from a steady increase in fishing mortality to $\mathrm{F}_{\text {ages 2-6 }}=0.6$ (ca. 45\%) in 1985 and a high bycatch of juveniles in the industrial fisheries for sprat. Following a period of four years of below average recruitment (year classes 1987-91) SSB fell rapidly to below 500,000 tonnes in 1993. Fishing mortality increased rapidly averaging $\mathrm{F}_{\text {ages } 2-6}=0.75$ (ca. $52 \%$ ) over the period 1992-95 and recorded landings regularly exceeded the TAC. The North Sea industrial fishery for sprat developed rapidly over this period with the annual catch increasing from 33,000 tonnes in 1987 to 357,000 tonnes by 1995 . With the $10 \%$ by-catch limit as the only control on the catch of immature herring, there was a consequent high mortality on juvenile herring which averaged $76 \%$ of the total catch in numbers of North Sea autumn spawners over this period.

During the summer of 1991 the presence of the parasitic fungus Ichthyophonus spp was noted in the North Sea herring stock. All the evidence suggested that the parasite was lethal to herring and that its occurrence could have a significant effect on natural mortality in the stock and ultimately on spawning stock biomass. High levels of infection were recorded in the northern North Sea north of latitude $60^{\circ} \mathrm{N}$ whilst infection rates in the southern North Sea and English Channel were very low. Efforts were made to estimate the prevalence of the disease in the stock through a programme of research vessel and commercial catch sampling. This led to estimates of annual mortality up to $16 \%$ (Anon., 1993) which was of the same order as the estimate of fishing mortality at the time. It was recognised that the behavioural changes and catchability of infected fish affected the reliability of the estimate of prevalence of the disease in the population. The uncertainty about the effect on stock size varied between estimates of $5 \%$ to $10 \%$ and $20 \%$. Continued monitoring of the progress of the disease showed that by 1994 the prevalence in the northern North Sea had fallen from $5 \%$ in 1992 to below $1 \%$ and confirmed that the infection did not appear to be spreading to younger fish. Ultimately it was concluded that the disease had caused high mortality in the northern North Sea during 1991 and subsequently declined to the point where by 1995 the disease induced increase in natural mortality was insignificant.

The increased fishing pressure during the first half of the 1990's and the disease induced increase in natural mortality led to serious concerns about the possibilities of a stock collapse similar to that in the late 1970 's. Reported landings continued at around 650,000 tonnes per year whilst the spawning stock began to decline again from over 1 million tonnes in 1990. The assessments at that time were providing an over optimistic perception of the size of the spawning stock and, for example, it was not until 1995 that it was realised that the SSB in 1993 had already fallen below 500,000 tonnes. This was well below the minimum biologically accepted level of 800,000 tonnes (MBAL) which had been set for this stock at that time.

## H. 3 Management and ACFM advice

In 1996, the total allowable catches (TACs) for Herring caught in the North Sea (ICES areas IV and Division VIId) were changed mid-year with the intention of reducing the fishing mortality by $50 \%$ for the adult part of the stock and by $75 \%$ for the juveniles. For 1997, the regulations were altered again to reduce the fishing mortality on the adult stock to 0.25 and for juveniles to less than 0.1 with the aim of rebuilding the SSB up to 1.1 million t in 1998.

According to the EU and Norway agreement adopted in December 1997, efforts should be made to maintain the SSB above the MBAL (Minimum Biologically Acceptable Level) of 800,000 tonnes. An SSB reference point of 1.3 million has been set above which the TACs will be based on an $F=0.25$ for adult herring and $F=0.12$ for juveniles. If the SSB falls below 1.3 million tonnes, other measures will be agreed and implemented taking account of scientific advice. The management agreement was revised in 2004 and now reads:

The stock is managed according to the EU-Norway Management agreement which was updated on 26 November 2004, the relevant parts of the text are included here for reference:

1. Every effort shall be made to maintain a level of Spawning Stock Biomass (SSB) greater than the 800,000 tonnes (Blim).
2. Where the SSB is estimated to be above 1.3 million tonnes the Parties agree to set quotas for the directed fishery and for by-catches in other fisheries, reflecting a fishing mortality rate of no more than 0.25 for 2 ringers and older and no more than 0.12 for 0-1 ringers.
3. Where the SSB is estimated to be below 1.3 million tonnes but above 800,000 tonnes, the Parties agree to set quotas for the direct fishery and for by-catches in other fisheries, reflecting a fishing mortality rate equal to:

$$
\begin{aligned}
& 0.25-(0.15 *(1,300,000-S S B) / 500,000) \text { for } 2 \text { ringers and older, and } \\
& 0.12-\left(0.08^{*}(1,300,000-S S B) / 500,000\right) \text { for } 0-1 \text { ringers. }
\end{aligned}
$$

4. Where the SSB is estimated to be below 800,000 tonnes the Parties agree to set quotas for the directed fishery and for by-catches in other fisheries, reflecting a fishing mortality rate of less than 0.1 for 2 ringers and older and less than 0.04 for 0-1ringers.
5. Where the rules in paragraphs 2 and 3 would lead to a TAC which deviates by more than $15 \%$ from the TAC of the preceding year the Parties shall fix a TAC that is no more than $15 \%$ greater or $15 \%$ less than the TAC of the preceding year.
6. Not withstanding paragraph 5 the Parties may, where considered appropriate, reduce the TAC by more than $15 \%$ compared to theTAC of the preceding year.
7. By-catches of herring may only be landed in ports where adequate sampling schemes to effectively monitor the landings have been set up. All catches landed shall be deducted from the respective quotas set, and the fisheries shall be stopped immediately in the event that the quotas are exhausted
8. The allocation of TAC for the directed fishery for herring shall be $29 \%$ to Norway and $71 \%$ to the Community. The by-catch quota for herring shall be allocated to the Community
9. A review of this arrangement shall take place no later than 31 December 2007.
10. This arrangement enters in to force on 1 January 2005.

## H. 4 Sampling of commercial catch

Sampling of commercial catch is conducted by the national institutes. HAWG has recommended for years that sampling of commercial catches should be improved for most of the stocks. In January 2002, a new directive for the collection of fisheries data was implemented for all EU member states (Commission Regulation 1639/2001). The provisions
in the "data directive" define specific sampling levels. As most of the nations participating in the fisheries on herring assessed here have to obey this data directive, the definitions applicable for herring and the area covered by HAWG are given below:

| Area | SAMPLING LEVEL PER 1000 t Catch |  |  |
| :--- | :--- | :--- | :--- |
| Baltic area (IIIa (S) and IIIb-c) | 1 sample of <br> which | 100 fish measured and | 50 aged |
| Skagerrak (IIIa (N)) | 1 sample | 100 fish measured | 100 <br> aged |
| North Sea (IV and VId): | 1 sample | 50 fish measured | 25 aged |
| NE Atlantic and Western Channel ICES areas II, V, <br> VI, VII (excluding d) VIII, IX, X, XII, XIV | 1 sample | 50 fish measured | 25 aged |

Exemptions to the above mentioned sampling rules are:
Concerning lengths:
(1) the national programme of a Member State can exclude the estimation of the length distribution of the landings for stocks for which TACs and quotas have been defined under the following conditions:
(i) the relevant quotas must correspond to less than $5 \%$ of the Community share of the TAC or
to less than 100 tonnes on average during the previous three years;
(ii) the sum of all quotas of Member States whose allocation is less than $5 \%$, must account for
less than $15 \%$ of the Community share of the TAC.
If the condition set out in point (i) is fulfilled, but not the condition set out in point (ii), the relevant Member States may set up a coordinated programme to achieve for their overall landings the implementation of the sampling scheme described above, or another sampling scheme, leading to the same precision.

## Concerning ages:

(1) the national programme of a Member State can exclude the estimation of the age distribution of the landings for stocks for which TACs and quotas have been defined under the following conditions:
(i) the relevant quotas correspond to less than $10 \%$ of the Community share of the TAC or to
less than 200 tonnes on average during the previous three years;
(ii) the sum of all quotas of Member States whose allocation is less than $10 \%$, accounts for less than $25 \%$ of the Community share of the TAC.

If the condition set out in point (i) is fulfilled, but not the condition set out in point (ii), the relevant Member States may set up a coordinated programme as mentioned for length sampling.

If appropriate, the national programme may be adjusted until 31 January of every year to take into account the exchange of quotas between Member States;

The HAWG reviewed the implementation of the new sampling regime for the EU countries in 2003. It was expected that the overall sampling level might be improved, and this was demonstrated e.g. for North Sea herring in 2002 and 2003. However, there is concern that the
new regime may lead to a deterioration of sampling quality, because it does not assure an appropriate sampling of different métiers (each combination of fleet/nation/area and quarter). Given the diversity of the fleets harvesting most stocks assessed by HAWG, an appropriate spread of sampling effort over the different métiers is more important to the quality of catch at age data than a sufficient overall sampling level. The EU data directive appears to not assure this. The WG therefore recommends that all metiers with substantial catch should be sampled (including by-catches in the industrial fisheries), that catches landed abroad should be sampled and information on these samples should be made available to the national laboratories.

## H. 5 Terminology

The WG uses "rings" rather than "age" or "winter rings" throughout the report to denominate the age of herring, with the intention to avoid confusion It should be observed that, for autumn spawning stocks, there is a difference of one year between "age" and "rings". HAWG in 1992 (ICES 1992/Assess:11) stated that
"The convention of defining herring age rings instead of years was introduced in various ICES working groups around 1970. The main argument to do so was the uncertainty about the racial identity of the herring in some areas. A herring with one winter ring is classified as 2-years-old if it is an autumn spawner, and one-year-old if it is a spring spawner. Recording the age of the herring in rings instead of in years allowed scientists to postpone the decision on year of birth until a later date when they might have obtained more information on the racial identity of the herring.

The use of winter rings in ICES working groups has introduced a certain amount of confusion and errors. In specifying the age of the herring, people always have to state explicitly whether they are talking about rings or years, and whether the herring are autumn- or spring spawners. These details tend to get lost in working group reports, which can make these reports confusing for outsiders, and even for herring experts themselves. As the age of all other fish species (and of herring in other parts of the world) is expressed in years, one could question the justification of treating West-European herring in a special way. Especially with the present trend towards multispecies assessment and integration of ICES working groups, there might be a case for a uniform system of age definition throughout all ICES working groups.

However, the change from rings to years would create a number of practical problems. Data files in national laboratories and at ICES would have to be adapted, which would involve extra costs and manpower. People that had not been aware of the change might be confused when comparing new data with data from old working group reports. Finally, in some areas (notably Division IIIa), the distinction between spring- and autumn spawners is still hard to make, and scientists preferred to continue using rings instead of years.

The Working Group discussed at length the various consequences of a change from rings to years. The majority of the Group felt that the advantages of such a change did not outweigh the disadvantages, and it was decided to stick to the present system for the time being."

The text table below gives an example for the correlation between age, rings and year class for the different spawning types in late 2002:

| Year CLASS (AUTUMN SPAWNERS) | $\mathbf{2 0 0 1 / 2 0 0 2}$ | $\mathbf{2 0 0 0 / 2 0 0 1}$ | $\mathbf{1 9 9 9 / 2 0 0 0}$ | $\mathbf{1 9 9 8 / 1 9 9 9}$ |
| :--- | :--- | :--- | :--- | :--- |
| Rings | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ |
| Age (autumn spawners) | 1 | 2 | 3 | 4 |
| Year class (spring spawners) | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 0}$ | $\mathbf{1 9 9 9}$ |
| Rings | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ |
| Age (spring spawners) | 0 | 1 | 2 | 3 |

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## Annex 4: TECHNICAL MINUTES

## REVIEW OF CELTIC SEA STOCKS

Reviewers: Pablo Abaunza, Spain (chair)<br>Tiit Raid, Estonia<br>Peter Lewy, Denmark<br>Claus Hagebro, from ICES Secretariat<br>Chair of Herring Assessment Working Group: Mark Dickey-Collas, Netherlands

## General:

The Celtic Sea’s review group dealt with the revision of the following herring stocks assessed in the Herring Assessment Working Group (HAWG):

- Celtic Sea and Division VIIj herring (benchmark assessment)
- Herring in Division VIa (North) (West of Scotland herring) (Update assessment)
- Herring in Divisions Via (South) and VIIb,c. (Update assessment)
- Irish Sea Herring [Division VIIa]. (Update assessment)

The "Clyde" herring stock has very poor data with very low research intensity (only catch data) and as a result no advice for this stock was presented. Therefore, the Clyde herring stock was not considered in the reviewing process.

The reviewers commend the thorough report and the quality of the work developed in relation to the herring stocks in the Celtic Seas. Although there was only one benchmark assessment the effort to improve the knowledge on the state of the stock was significant in all the reviewed herring stocks (e.g. exploring with new assessment methods like Bayesian approaches or revising historical catch data).

The Review Group acknowledged the work made by HAWG in relation to identifying possible regime shifts in herring productivity and in exploring the relationship between these changes and the environmental variability. The reviewers consider that this is a very useful approach to understand the dynamics of the herring stocks and that it is in tune with the ecosystem approach. In this sense the reviewers would also like to support the concerns stated by the chair of the HAWG about some inaccuracies in the descriptions-recommendations on Celtic Seas raised by WGRED (Working Group on Regional Ecosystem Definition).

## Assessment of Celtic Sea and Division VIIj herring (benchmark).

Stock identity has already been considered after the results provided from the EU-funded WESTHER project.

This stock has relatively good information on catches, and a high precision in catch at age data. However, the stock show a high variability in its dynamics which is also reflected in the tuning acoustic survey used for the assessment, which is very noisy. Therefore, there is no possibility to obtain precise estimates of the current SSB and F. Exploratory data analysis on cohort catch curves was carried out and various exploratory assessments are also presented, including Bayesian methods to account for a more realistic estimate of the uncertainty, simpler models (CSA), VPA-based methods (XSA) and statistical catch at age-like (ICA) in which
different settings were explored. These assessments showed a similar declining trend for SSB in recent years.

The reviewers agreed with the WG conclusion that as there is no agreed final assessment, the basis for the catch advice is limited to overall trends qualitative assessment results rather than year specific estimates of mortality or biomass.

Catch tables (4.1.3.1 and 4.1.3.2) should be cross-checked for possible errors in the discard estimates.

The acoustic survey plays a key role in the assessment of this stock, since at this moment it is the only valuable information for tuning the assessment. The reviewers consider that a summary of the revision of the acoustic series made last year should also be included in the text, in addition to a more clear explanation of the survey design (i.e. if it is adaptive or not, changes in the timing in which the survey is carried out, etc.).

In relation to the assessment, catch residual plots indicate model problems with clear year and age effects. This could be due to the problems with the assumptions of separability. It is therefore important to investigate why this is the case for instance by investigating if the selection pattern changes over years. This could be done by using a model allowing for such changes.

In the report there is an extensive description of the strength of the recent year classes. However, the reviewers considered that compared with the historical series there is no any strong year class in the last decade (at most only of medium strength). On the other hand, it seems that there is an inconsistency between the recruitment series showed in figure 4.6.2.4 (a) and figure 4.10 .1 (c). In figure 4.6.2.4 there is one point more. This makes the identification of particular cohorts more difficult, and should be checked changing also the text accordingly where appropriate.

The figure on the residual patterns around the stock recruitment model show clearly the absence of trends but the classical plot of stock-recruitment relationship should also be included to complement this information.

## Herring in Division Via (North) (West of Scotland herring) (update)

Stock identity has already been considered after the results provided from the EU-funded WESTHER project. Catches have been revised with respect to reallocation and misreporting and sampling of catch improved in 2006. This new revision of historical data required exploration. An exploratory assessment (ICA) using the same settings as last year but using the new data set was carried out to look at possible changes in the perception of the state of the stock. The reviewers consider that it should be helpful to understand the impact of these data changes if a figure comparing the historical series of SSB and F between the assessment with the revised data and with the "old" data is included.

The inclusion or exclusion in the assessment of two low values in the historical series of the tuning survey was also explored and the differences were within the bounds of the confidence intervals of the assessment. The retrospective pattern supports the perception of a noisy but fairly well balanced assessment. The assessment showed a decreasing trend in SSB and that the stock was heavily exploited in 2006.

Deterministic short-term projections and yield-per-recruit analysis were also carried out using the standard software. The proposed management plan has also been included in the projections.

The reviewers agreed with last year's conclusion that the SSB estimate is uncertain.

## Irish Sea herring (Division VIIa (North)) (update)

Stock identity has already been considered after the results provided from the EU- funded WESTHER project

An exploratory assessment using the same settings as last year and two tuning surveys was carried out but no analytical assessment is presented. However, the chair of the HAWG informed that some of the input data corresponding to the last year of the exploratory assessment were wrong. This information was obtained after the HAWG a few weeks after the meeting and consequently it does not appear in the report. The state of the stock is therefore unknown but taking in consideration only the information from surveys it is likely that the SSB is stable at low level. The reviewers recommend checking the input data for the exploratory assessment as soon as possible. A contraction of the age structure is described in catches and from surveys information. Catches have also been low in recent years and fishing activity has not varied considerably. The HAWG considered that the maintenance of recent catches should not be detrimental to the stock.

The reviewers agreed with the WG conclusion that since there is no agreed final assessment, the basis for the catch advice is limited to overall trends qualitative assessment results rather than year specific estimates of mortality or biomass

## Herring in Division VIa (South) and VIIb,c (update)

Stock identity has already been considered after the results provided from the EU-funded WESTHER project.

There is an exploratory assessment available for this stock based on a separable VPA. Although there is an abundance acoustic estimate since 5 years ago, the historical series is too short to consider it as a tuning survey in an analytical assessment. The reviewers would like to support the continuity of this survey series that could be used in the future assessments of this stock. In absence of an agreed assessment, it was not considered to carry out any predictions. The SSB estimates from this non-tuned assessment are uncertain but it is likely to be at historical low level in recent years. The figure 6.6.2.3 (b), shows the historical series of SSB, but the level of $\mathrm{B}_{\mathrm{lim}}$ is erroneous and should be changed to the actual level of 81000 t .

The reviewers agreed with the WG conclusion that this stock should be exploited with great caution and with that there is no agreed final assessment. The basis for the catch advice is limited to overall trends qualitative assessment results rather than year specific estimates of mortality or biomass

## Annex 5: Technical Minutes

# REVIEW OF NORTH SEA STOCKS 

Participants<br>Reidar Toresen (chair)<br>Eero Aro<br>Steve Cadrin<br>Andre Forest<br>Morten Vinther<br>Wg chairs: Mark Dickey-Collas (HAWG), Chris Darby (WGNSSK)

## Stocks assessed by the HAWG

## Herring in IV

## General comments:

An update assessment was completed. The stock is declining because of poor recruitment in recent 5 years.

There is some misreporting of catches by area and overshoot of landings. Some of the catch data has been revised. Some input data was adjusted.

## Technical comments:

The RG questioned why the 3,4 and 5 year old estimates of the IBTS surveys were used in the assessments.

The best estimate of the older ages is in the catch, but the RG questioned the quality of the catch data and sampling of catches.

The RG also had several questions about the surveys: Why does the IBTS survey perform badly, for age group 2? Why doesn't the acoustic survey pick up older ages? Why are there old fish in the catch and not in the surveys?

The RG also noted the inconsistent accounting of small-mesh herring fisheries, because the Norwegian catches of small herring are included in the A-fleet.

A minor statistical note is that Q-Q plots may not be optimal for such low sample sizes. When sample size is small ( $\mathrm{n}<50$ ), q-q plots are sensitive to the number of samples, particularly in the tails of the distribution. A more appropriate evaluation of normality for low sample sizes is ranked normal deviates, rather than quantiles (Sokal \& Rohlf 1995). R code for the procedure:

## rankit<- qnorm(ppoints(n))[order(b=variable)]

qqplot (variable, rankit)\# makes quantile -quantile plot
xy1 <-qqplot (variable, rankit, plot=FALSE)\# makes files with scores, normal quantiles
r1<- lsfit (xy1\$x, xy1\$y) \# makes linear fit
abline (r1, lty=2, col="blue") \# plots line based on intercept, slope

## Conclusions:

The assessment was accepted.
SSB is under $\mathrm{B}_{\text {trig }}$ but is declining.
F is greater than target F (0.35). The management rule is not robust to implementation error. Last year managers agreed on a TAC that was greater than that indicated by the management rule.

A mistake in input data file was found this year. This only had a small impact on the assessment (less than 1\%). The RG decided to use the present assessment, but the mistake needs to be addressed and corrected as basis for next year's assessment of the stock.

The large overshoot of F is explained with the fact that managers have agreed on too high TACs in recent years. This should be interpreted as implementation error of the Management Plan.

SSB is below $B_{\text {trig }}$ of 1,3 million tones and decreasing.
$F$ is above target $F$ of 0,25 , and increasing.
Recruitment is very low and has been low for 5 successive years.

## Sprat in IV

## General comments:

An experimental assessment was completed. Data from IBTS and a short time series of acoustic data is available. The assessment has never been used for advice. A regression between IBTS index and catch the following year has been used as catch forecast.

## Technical comments:

The RG recommends that the WG try to develop the acoustic abundance estimates, and see if these could be used in assessments.

The RG proposes that a data transformation of the IBTS survey data should be explored to reduce the effect of a few large tows in the IBTS surveys.

## Conclusions:

There is no evidence that the catch levels have created problems for the stock. An in-year recommendation should be made on the same basis as the recommendation made last year. However, the RG noted that if three points in the regression are removed (1989, 1994 and 1995) associated with a large-tow effect and years when there is a lot of herring in the catches (1994-1995), the regression may be more reliable?

## Western Baltic spring spawners. Herring in Illa

## General comment:

An update assessment was completed. The assessment is noisy, with large residuals, and huge year-effects. Each source of input data covers only a portion of the stock.

## Technical comments:

The RG noted that input data are all weighed 1 , and recommended using the same weighing process for the data for this stock as for North Sea herring.

The RG concluded that the assessment is not reliable for status determination, because it lumps together information on different parts of the stock. Retrospective pattern for recruits are very bad. Residuals are large, and there are year-effects on the residuals of the assessment. The RG felt that the quality of the assessment is poor. The estimate of 0 - group the most recent year is particularly bad.

The RG recommends that there should be a survey which covers all components of the stock.

## Conclusions:

The assessment was accepted, with a recommendation that there should be a benchmark assessment next year.

SSB cannot be evaluated in relation to reference values because they are not defined, but SSB seems to have stabilized.

Fishing mortality seems to have stabilized at levels around 0,5 , which is rather high for a herring stock.

There are signs of a declining trend in recruitment

## SPRAT in Illa

Landings since 1974.


[^0]:    ${ }^{4}$ Including IVa East
    ${ }^{5}$ Negative unallocated catches due to misreporting from other areas
    ${ }^{6}$ Altered in 2000 on the basis of a Bayesian assessment on m isreporting into IVa (North)
    ${ }^{7}$ Including any by-catches in the industrial fishery
    ${ }^{8}$ May include misreported catch from VIaN and discards
    ${ }^{9}$ Figure altered in 2001
    ${ }^{10}$ Including 1057 t of local spring spawners
    ${ }^{11}$ Figures verified and altered if needed in 2003 by SG Rednose (ICES 2003/ACFM:10)

[^1]:    ${ }^{2}$ Catches of Norwegian spring spawning herring removed (taken under a separate TAC)
    ${ }^{3}$ Included in IVa West
    ${ }^{4}$ Negative unallocated catches due to misreporting into other areas
    ${ }^{5}$ Including any by-catches in the industrial fishery
    ${ }^{6}$ These catches (including some fjord-type spring spawners) are taken by Norway under a separate quota south of $62^{\circ} \mathrm{N}$ and are not included in the Norwegian North Sea catch figure fir this area

[^2]:    ${ }^{2}$ Discards partly included in unallocated landings
    ${ }^{3}$ Negative unallocated catches due to misreporting from other areas
    ${ }^{4}$ Including any by-catches in the industrial fishery
    ${ }^{5}$ May include discards. Negative unallocated due to misreporting into other areas
    ${ }^{6}$ Figures verified and altered if needed in 2003 by SG Rednose (ICES 2003/ACFM:10)
    ${ }^{14}$ Figure altered in 2004

[^3]:    ${ }^{2}$ Landings from the Thames estuary area are included in the North Sea catch figure for UK (England)
    ${ }^{3}$ Discards partly included in unallocated landings
    ${ }^{4}$ May include misreported catch and discards
    ${ }^{9}$ Figures verified and altered if needed in 2003 by SG Rednose (ICES 2003/ACFM:10)
    ${ }^{10}$ Figure altered in 2002 (was 7851 t higher before)
    ${ }^{11}$ Thames/Blackwater herring landings: 107 t , others included in the catch figure for The Netherlands
    ${ }^{14}$ Figure altered in 2004

[^4]:    *) "Roundfish areas" are shown in the IBTS Manual (Add. ICES CM 2002/D:03)

[^5]:    Data for 1995 to 2001 was revised in 2003.

[^6]:    Corrections for the years 1991-1998 was made in WG2001, but are NOT included in the North Sea assessment.

[^7]:    Input units are thousands and kg - output in tonnes

[^8]:    Western Baltic Herring. Optimization A. $\mathrm{B}_{\mathrm{pa}}=$ lower SSB limit.

[^9]:    \# The 1997 survey is not on the same basis as the other years, it was conducted in June (all other surveys were carried out in July).

[^10]:    * no information, but catch is likely to be negligible

[^11]:    * Average for the preceding nine years

