# ICES HAWG REPORT 2009 

# Report of the Herring Assessment Working Group for the Area South of 62 N 

17-25 March 2009

ICES Headquarters, Copenhagen

# International Council for the Exploration of the Sea Conseil International pour l'Exploration de la Mer 

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## Executive Summary

The ICES herring assessment working group (HAWG) met for 7 days in March 2009 to assess the status of 7 herring stocks and 3 sprat stocks. The working group conducted update assessments for North Sea, Western Baltic and West of Scotland herring stocks and for North Sea sprat. Moreover, the group performed an assessment for Celtic Sea and Division VIIj herring which is presented as an update.

The SSB of North Sea autumn spawning herring in autumn 2008 was estimated at 1.0 million $t$, and is expected to remain below $B_{p a}(1.3$ million $t)$ in 2009. $F_{2-6}$ in 2008 was estimated at 0.24 , above the target $\mathrm{F}_{2-6}$ of 0.14 . The year classes since 2002 are estimated to be among the weakest since the late 1970s. Best estimates of catches in 2008 were 257900 t , a decrease from 406900 t in 2007. The Western Baltic spring spawning stock's SSB has been rather stable over the last decade, although the most recent value is in the lower quartile of all observations. Fishing mortality has also been stable in the same period but is larger than any proxy of $\mathrm{F}_{\text {msy. }}$. Recruitment has declined consistently since 2003 and the estimated number of 0-ringers in 2008 is the lowest observed value. The 2003 year class year class has been the largest component of the SSB for the last three years and has supported the stock during this period. However, this year class will pass out of the stock in the next two years, whilst its place will be taken by the sequence of poor year classes: a continuation of the decline in SSB can therefore be expected in the short and medium term. The Celtic Sea autumn and winter spawning stock has increased in size and SSB is now above $B_{p a}$ and mean $F_{2-5}$ has declined to the lowest estimate observed. Catch in 2008/2009 decreased to lowest in the series ( 5800 t ). Two strong and two weak year classes have recruited recently. In recent years the assessment was considered as indicative of trends. In 2009 HAWG puts forward an analytical assessment for this stock, which is considered by the WG to be stable enough to provide the basis for advice. West of Scotland autumn spanwing stock's SSB (in 2009) is 1.8 times Blim. The stock is currently fluctuating at a low level and is being exploited close to $\mathrm{F}_{\mathrm{msy}}$. Recruitment has been low since 1998. Catch in 2008 was 16000 t , a decrease from 29000 t in 2007. The WG evaluated the recently agreed management plan for this stock (slightly changed from the proposed plan) and found no substantive differences from the earlier evaluations of medium term risks of SSB<Blim, indicating that advice could be based on the agreed plan. West of Ireland (Division VIaS and VIIb,c) autumn- and winter/spring-spawning stock cannot be assessed analytically because no tuning data are available. However, there are indications that the stock is at a historically low level. Though current levels of SSB and F are not precisely known, there are no sign of stock recovery. Catch in 2008 was 13300 t , a decrease from 18000 t in 2007. Irish Sea autumn spawning herring SSB has been relatively stable for the last 10 years, and fishing mortality does not appear to be increasing above the recent average. Catches (4900 t in 2008) have been close to TAC level in recent years and the main fishing activity has not varied considerably. An increase in effort on the Mourne spawning component has been noted in the past three years. There is some evidence of increased recruitment in the stock in most recent years. Catches of the Clyde spring spawning stock were 676 t in 2008, but no information is available to perform an assessment.

Survey trends indicate that the stock size of North Sea Sprat has varied around an average level with no trend. There is no analytical assessment for this stock. The recruits account for a large proportion of the stock, and the fishery in a given year is very dependent on that year's incoming year class. The state of the stock is uncertain, and catches in 2008 were 61100 t , declining from 81000 t in 2007. The new data avail-
able for sprat in Division IIIa were too sparse to perform an assessment. The total landings decreased from 15700 t in 2007 to 9100 t in 2008. Sprat in VIId,e catch was somewhat higher than the recent average ( 3300 t in 2008). No assessment of this stock was possible.

The group answered one ad hoc request, for updated advice for western Baltic spring spawning herring

The working group also commented on the quality and availability of data, the problems with estimating the amounts of discarded fish, the use of the data system INTERCATCH, changes in mean weights of the stocks considered by the group and recent meetings and reports of relevance to HAWG.

## 1 Introduction

### 1.1 Participants

| Steven Beggs | UK/Northern Ireland |
| :--- | :--- |
| Stijn Bierman | The Netherlands |
| Massimiliano Cardinale | Sweden |
| Maurice Clarke (Co-Chair) | Ireland |
| Lotte Worsøe Clausen | Denmark |
| Mark Dickey-Collas | The Netherlands |
| Afra Egan | Ireland |
| Christina Frisk | Denmark |
| Tomas Gröhsler (Co-Chair) | Germany |
| Joachim Gröger | Germany |
| Clementine Harma | Ireland |
| Niels Hintzen | The Netherlands |
| Cecilie Kvamme | Norway |
| Henrik Mosegaard | Denmark |
| Peter Munk | Denmark |
| Mark Payne | Denmark |
| Beatriz Roel | UK/England \& Wales |
| Norbert Rohlf | Germany |
| Barbara Schoute | ICES Secretariat |
| John Simmonds | UK/Scotland |
| Dankert Skagen | Norway |
| Else Torstensen | Norway |
| Yves Verin | France |

Contact details for each participant are given in Annex 1.

### 1.2 Terms of Reference

2008/2/ACOM03 The Herring Assessment Working Group for the Area South of $\mathbf{6 2} \mathbf{2}^{\mathbf{N}}$ [HAWG] (Co-Chairs: Tomas Gröhsler, Germany and Maurice Clarke, Ireland) will meet at ICES Headquarters, 17-25 March 2009 to:
a ) compile the catch data of North Sea and Western Baltic herring on 17-18 March
b ) address generic ToRs for Fish Stock Assessment Working Groups 19-25 March (see table below).

The assessments will be carried out on the basis of the stock annex in National Laboratories, prior to the meeting. This will be coordinated as indicated in the table below.

| Fish <br> Stock | Stock Name | Stock Co- <br> ord. | Assesss. <br> Coord. 1 | Assess. <br> Coord. 1 | Advice |
| :--- | :--- | :--- | :--- | :--- | :--- |
| her-3a22 | Herring in Division IIIa and Subdivi- <br> sions 22-24 (Western Baltic Spring <br> spawners) | Denmark | Germany | Denmark | Advice |
| her-47d3 | Herring in Subarea IV and Division IIIa <br> and VIId (North Sea Autumn spawners) | Germany | NL | UK (Scot- <br> land) | Advice <br> her-irlsHerring in Division VIIa South of 52 <br>  <br> $30^{\prime}$ N and VIIg,h,j,k (Celtic Sea and |
| Ireland | Ireland |  | Same <br> advice <br> as last |  |  |

\(\left.$$
\begin{array}{|l|l|l|l|l|l|}\hline & \text { South of Ireland) } & & & \begin{array}{l}\text { year } \\
\hline \text { her-irlw } \\
\hline \text { her-nirs } \\
\text { VIIb,c }\end{array} & \begin{array}{l}\text { Herring in Divisions VIa (South) and } \\
30^{\prime} \text { N (Irish Sea) }\end{array}
$$ <br>
\hline advice <br>
as last <br>

year\end{array}\right]\)| Ireland |
| :--- |
| her-vian |
| Herring in Division VIa (North) |
| Spr-nsea |

HAWG will report by 30 March 2009 for the attention of ACOM.

### 1.3 Working Group's response to ad hoc requests

### 1.3.1 EU Baltic Pelagic HCR Special request

a) For WBSS herring advice provided in 2008 on the request EC-DG FISH 27.03.2007-021 should be updated as necessary.
b) Following recommendations in this advice, results from the benchmark assessment carried out in March 2008 should be taken into account and advice management options be developed in light of the multi-fleet fisheries on this stock.
c) In addition, mixing with NSAS herring in IIIa should be taken into account
d) Advice should be provided for a fixed allocation of catch options between Div. IIIa and SD 22-24.

## Response

## Advice for 2010

- An option realizing the advised HCR in 2008 has been included as a scenario in the Western Baltic spring spawning herring (WBSS) advice for 2010 according to the following rule:
- Target fishing mortality is 0.25 ,
- This results in a TAC (IIIa + SD 22-24) change larger than $15 \%$, thus the TAC change was restricted to $+15 \%$ (** value ${ }^{* *}$ ). The resulting SSB in 2011 is indicated to be below Btrig [C] 110000 t therefore this option was not used as a basis for advice.
- Results from the benchmark assessment carried out in March 2008 have been taken into account in the advice. Like earlier years, advice options are routinely given in light of the multi-fleet fisheries:

| North Sea | Fleet A | Directed herring fisheries with purse-seiners and trawlers. <br> Bycatches in industrial fisheries by Norway are included. |
| :--- | :--- | :--- |
|  | Fleet B | Herring taken as bycatch under EU regulations. |
| Division IIIa | Fleet C | Directed herring fisheries with purse-seiners and trawlers |
|  | Fleet D | Bycatches of herring caught in the small-mesh fisheries |
| Subdivision 22- <br> 24 | Fleet F | All herring fisheries in Subdivisions 22-24 |

- In addition, mixing with the North Sea Autumn Spawning (NSAS) herring in IIIa is routinely taken into account. For 2009, the expected catch of WBSS in IIIa was calculated assuming the same WBSS proportions in the catch of each fleet in 2009 as that in 2008 neglecting the small amount of about 120 t WBSS taken in Division IVaE by the A-fleet.


## Allocating catches between areas

- The ICES advice for the area is used in a TAC management system dividing Division IIIa from SD 22-24. The ICES short-term forecast assumes a ratio of catches over the two management areas based on recent landings. For the 2010 advice, based on the 2008 catches, a 47 : 53 ratio between IIIa : SD 22-24 catches is used. The historical allocation pattern is close to $50 / 50$ between Division IIIa and SD 22-24 as a result of informal allocation.

Considering the spatial distribution of the different life-stages over the seasons, ICES recommends that a specific rule be incorporated in the management plan for WBSS determining a fixed allocation. Other way round...The allocation of the TACs between the two areas should match the one used in the ICES forecast... Using a different allocation corrupts the power of the prediction of the forecast.

## Source of information

ICES 2008. Report of the Workshop on Herring Management Plans (WKHMP), 4-8 February, ICES Headquarters Copenhagen. ICES CM 2008/ACOM: 27. 2 pp.

### 1.4 Reviews of groups or work important for the WG

HAWG was briefed throughout the meeting about other groups and projects that were of relevance to their work. Some of these briefings and/or groups are described below.

### 1.4.1 Meeting of the Chairs of Assessment Related Expert Groups [WGCHAIRS]

HAWG was informed about the WGCHAIRS meeting in January 2009. The presentation focused on the following main outcome relevant for HAWG:

Reviews of EG reports: There is still a need to clarify the role of reviewers for stock assessment updates. The reviewers only need to determine that the assessment follows the rules given in the stock annex. The reviews should not act like a benchmark. There is a need to "caveat" reviews that are appended to WG reports to indicate that they are constructive feedback from peers, but that they are not necessarily superior to the Expert Group in terms of insights or validity.

Draft Advice: According to Council decision, it is the responsibility of ADGs to draft advice in order to facilitate integration. However, EG ToRs should be constructed to respond to requests for advice. If EGs prepare concise and focused text, tabular material, and figures that respond to ToRs, their work will have a clear impact on advice. Information like: Impact of fisheries on the ecosystems, Regulations and their effects, Changes in fishing technology and fishing patterns, Impact of the environment on the fish stock etc. should only be given in the advice, when any of this information was used in the assessment. A new SUBFORMAT for 2010 or later is proposed and under discussion.

Enhanced Integration and the ecosystem approach: Benchmark workshops are a mechanism to enhance integration and the ecosystem approach. EGs are asked to provide a list of stocks to be benchmarked in 2010, including a list of the criteria set up by the Benchmark Workshop Planning Group (PGBWK) (ICES CM 2008/ACOM:62). Regional ecosystem descriptions should be maintained even if this needs to be done ad hoc in 2009. They should be taken into account in future advice to the extent this is scientifically justified.

Working documents: An archiving system for these documents would be good, this works on SharePoint. The ICES Secretariat is looking at a way to keep up a longer time period for availability of these documents.
Planning Group on Commercial Catch, Discards and Biological Sampling (PGCCDBS): HAWG is asked to nominate one contact person, which should

- compile data issues addressed during our meeting that could possibly compromise the quality of the assessment,
- be responsible for communicating this message to relevant groups that deal with data collection (RCM \& PGCCDBS) personally or through the ICES secretariat.

This would also mean that the nominated member should/has to participate in all relevant data collection meetings (RCM \& PGCCDBS).

### 1.4.2 Planning Group of International Pelagic Surveys [PGIPS]

The Planning Group for Pelagic Surveys (PGIPS, formerly PGHERS) has met in January 2009 (ICES 2009/LRC:02) to co-ordinate acoustic and larvae surveys in the North Sea, the Malin Shelf and the Western Baltic; to combine recent survey results for assessment purposes and to elucidate parameters influencing these calculations.
Review of larvae surveys in 2008/2009: Six survey metiers were covered in the North Sea. Larvae abundance has increased in all observed areas, with the exception of the Buchan area. The Multiplicative Larval Abundance Index indicates that the SSB has increased compared to last year.
Workshop on the Identification of clupeid larvae (WKIDCL): A workshop should take place in Hamburg, Germany, from 1-3 September 2009 to review available information on the identification of clupeid fish larvae; to identify sources of misidentification of clupeid larvae and to establish an agreed identification key for participants in clupeid larvae surveys, e.g. for the IHLS in the North Sea, the Irish Sea, the IBTS (MIK index), the Rügen HLS and the Norwegian Spring Spawning herring larvae surveys.

North Sea, West of Scotland and Malin Shelf summer acoustic surveys in 2008: Eight acoustic surveys were carried out during late June and July 2008 covering the

North Sea, West of Scotland and the Malin Shelf area. The estimate of North Sea autumn spawning herring spawning stock is at 1.8 million tonnes. This is slightly higher than the previous year ( 1.2 million tonnes). The West of Scotland estimates of SSB are 788000 tonnes. This is the second highest estimate in the time series. The survey did not detect many immature fish this year.
For the first time, a synoptic survey of what is currently considered the Malin Shelf population of herring was carried out. This provided an estimate comprising four stocks to the west of the British Isles: the West of Scotland herring stock in Division VIaN; the Clyde stock; the stock in Division VIaS and VIIb, c and the Irish Sea stock. The Malin Shelf estimate of SSB was 826000 tonnes and is largely dominated by the west of Scotland estimate.

Sprat: In most recent years, there is a downward trend in North Sea sprat. In 2008, the total biomass was estimates to 270000 tonnes, which is a reduction by $25 \%$ when compared to last year. The majority of the stock consists of mature fish. The sprat stock is dominated by 1- and 2-year old fish representing more than $95 \%$ of the biomass.

In Division IIIa, sprat was abundant in the Kattegat only. No sprat was observed in the Skagerrak area. The biomass has significantly decreased to 12000 tonnes.

Western Baltic acoustic surveys in 2008: A joint German-Danish acoustic survey was carried in the Western Baltic in October 2008. The estimate of Western Baltic spring spawning herring is about 124000 tonnes in Subdivisions 22-24 and is dominated by young herring as in former years. The present overall estimates are low both in terms of abundance and biomass, when compared to the long term mean. The estimated total sprat stock is around 60000 tonnes and indications are found for a weak upcoming year class.

### 1.4.3 Work on Multi-annual Management of pelagic stocks in the Baltic [WKMAMPEL]

The ICES Workshop on Multi-annual management of Pelagic Fish Stocks in the Baltic [WKMAMPEL] met in February 2009 in response to an EC request to develop a multi-annual plan for the management of the pelagic fish stocks in the Baltic Sea. The Western Baltic Spring Spawning, and specifically the fishery on this stock in SD 22-24 was considered as part of this process.

Management plans for this stock have previously been examined by the WKHMP working group in February 2008. These simulations suggested that a target F ("A" parameter in a harvest control rule) should be set no higher than 0.25 . Exploration of different juvenile selection patterns indicated that at high fishing mortalities the proportion of simulations falling below Bim increased with increasing juvenile selection. Limitations on the year-to-year variation in TAC (" $B$ " parameter) were recommended to be $15 \%$. WKHMP was unable to make a specific recommendation on the level of the trigger biomass ("C" parameter).

These evaluations were based on population parameters from the 2007 stock assessment. The stock was the subject of a benchmark assessment in March 2008, which improved the quality of the assessment, although the overall perception of the development of the stock changed little. Work performed by ICES in August 2008 updated the results of WKHMP to incorporate the results of the benchmark assessment. The new simulations differed little from those presented in WKHMP and ICES concluded
that the conclusions drawn by WKHMP were not changed by the benchmark assessment process.
Both WKHMP and WKMAMPEL highlighted several concerns about the appropriateness of the tools employed to assess the proposed management plans, noting the inherent complexity of the Western Baltic Spring Spawning population and associated fishery. WKHMP and WKMAMPEL recommended that further work be performed to develop tools appropriate for this system. Work is currently ongoing, through the "JAKFISH" and "GAP" EU projects, to develop such tools. This work is expected to run for the duration of the JAKFISH project (until 2011). Additionally, careful scientific evaluations by the ICES quality assurance system and by the ICES Advisory Committee will be required. It is therefore expected that it will take at least 1-2 years before such work could be completed and implemented.

### 1.4.4 Linking Herring 2009 [ICES/PICES/GLOBEC sponsored symposium]

The Linking Herring symposium was organized to link our understanding of herring biology, population dynamics and exploitation in the context of ecosystem complexity. It is beyond argument that herring play a pivotal role in shaping the structure and dynamics of many boreal continental-shelf ecosystems. Thus, in moving to an ecosystem approach to fisheries management, the time seemed right for ICES to hold another herring symposium. Since the last ICES symposia on herring in the 1960s (ICES Herring Symposium, 1961; Biology of Early Stages and Recruitment Mechanisms of Herring, 1968), many of the former paradigms have been rejected and substantial progress has been made by striking out along new avenues. The symposium covered new research from both the ICES and PICES community.

The symposium took place from the 26th to the 29th August 2008, at the National University of Ireland, Galway, Ireland. The conference was co-sponsored by Marine Institute (Ireland), Institute of Marine Research (Norway), ICES, the Irish Tourist Board, PICES and Wageningen IMARES (The Netherlands) and supported by GLOBEC. In total there were 80 presentations, 64 oral and 16 posters. These studied the Atlantic (NE and NW), Pacific (NE and NW), Baltic and Arctic herrings. Delegates, numbering 100 in total, attended from Ireland, UK, Norway, Denmark, Italy, France, the Netherlands, Germany, Canada, USA, Russia, Latvia, Iceland and Poland. The local organisation was lead by Maurice Clarke (Marine Institute) and Patricia Walsh (National University of Ireland, Galway).
According to expectations, Linking Herring was an exciting symposium that successfully described the state of the art in herring science and management. However, it showed that there are still huge challenges ahead, particularly in understanding the role of herring within the ecosystem approach and how to translate this into actual management measures. With herring, fixed rules appear to be few, and any current paradigm is likely to shift in future. Exploiting herring in a sustainable manner may never be possible as its populations naturally come and go, even without exploitation. The example of Norwegian spring-spawning herring shows us that the choices of individuals belonging to a highly plastic species results in populations that adapt and vary over time. Our most important task is to ensure that any assumptions underlying the management advice reflect this feature of plasticity, even if we don't understand its genetic and phenotypic origin completely.

### 1.4.5 Study Group on the Evaluation of Assessment and Management Strategies of the Western Herring Stocks [SGHERWAY]

The ICES Study Group on the evaluation of assessment and management strategies of the western herring stocks [SGHERWAY] met in early December 2008 to consider issues surrounding the assessment and management of the herring stocks to the west of the British Isles.

SGHERWAY arose out of the EU funded project WESTHER which evaluated the uncertain stock identity of herring stocks to the west of the British Isles. Its results suggested a rearrangement of the stocks as they are currently assessed. SGHERWAY recognises the need to provide sound management advice for the western herring areas, and in particular the importance of ensuring as far as possible that there is no depletion of local components. Currently it is unclear what management regime would provide the most cost effective method for successful management and what data would be needed to support this management.

SGHERWAY considered that it is necessary to move towards management for this area through a series of iterations involving the following steps: (I) Investigation of a combined assessment of the three currently assessed stocks, VIaN, VIaS/VIIbc and VIIaN (to be called the Malin Shelf stock), including an investigation of the utility of a combined acoustic survey; (II) Examination of alternative management strategies based on their ability to deliver protection to local populations and provide cost effective information applicable for management of the new proposed stock unit of herring to the west of the British Isles (Malin Shelf); (III) Amendment of existing, or development of new, cost effective assessment and data collection schemes which will be required to support this management.

In December 2008, SGHERWAY was able to address the first of the two steps above. During the meeting the majority of the data required to perform a combined assessment of the three herring stocks were compiled and a combined assessment carried out using FLICA with the VIaN survey as the tuning index. This combined assessment gave a lower catchability than the current VIaN assessment, suggesting that the inclusion of additional catch from the VIaS/VIIb c and VIIaN stocks was an improvement. However, the retrospective pattern was very poor. This may be as a result of the partial coverage of the single tuning index used. In most years this survey does not extend as far as VIaS/VIIbc or VIIaN. Another possibility is that the selection pattern assumed for the fishery may not represent the combined fishery. The development of a time series of a synoptic acoustic survey of the Malin and Hebrides shelf areas will enable survey coverage to be extended to the whole sea area in which mixing of the various western herring stocks is thought to occur, and a more apposite tuning index to be developed. The first such synoptic survey was carried out in 2008. The area was surveyed in June/July 2008 by vessels from Scotland, Northern Ireland and the Republic of Ireland. The three survey estimates were combined in the same manner as the surveys in the North Sea. The Malin Shelf estimate of SSB was 826,000 tonnes and 4,007 million fish. This is largely dominated by the VIaN estimate.

A previously defined model was available to suffice as a good starting point to evaluate alternative management strategies for the metapopulations west of the British Isles. The simulations were run under a number of F and mixing settings. The model supplied allowed the study of some aspects of the dynamics of a fishery operating on mixed populations. In particular, the model allowed investigation of the effects of a sudden increase in catchability in one of the populations. However, shifts in fishing effort which should have an impact on all populations caught were not inves-
tigated. To be fully able to evaluate alternative management strategies for herring stocks west of the British Isles, it is recommended to adapt the model setup to coincide more with reality. Therefore, the model should be adjusted to a year-by-year evaluation platform, where management rules can be incorporated and fed back into the biological part of the model as well.

### 1.4.6 Planning Group on Commercial Catch, Discards and biological Sampling [PGCCDBS]

## Contact persons as link between AWG and PGCCDBS

PGCCDBS considered that the system of contact persons providing a link between ICES stock assessment Working Groups and PGCCDBS was insufficiently developed in 2008 to evaluate the success of this initiative. Furthermore, there did not appear to be a well-defined protocol for contacts officers to provide feedback from AWGs (assessment working groups). The PGCCDBS defined a suitable contact person profile to be:

- An active member of the relevant assessment group and the benchmark WKs related to the AWG stocks
- A participant of PGCCDBS or close contact with an attendee of that group.
- A participant of relevant regional coordinating meeting (RCM) or close contact with attendee of that group.

In order for the contact person to function effectively, PGCCDBS envisaged that the role should include the following tasks;

- Contact all stock coordinators (and assessors) that the HAWG represents in order to identify issues relevant to PGCCDBS.
- Ensure that all issues relevant to PGCCDBS and RCM's are entered in the table - "Stock Data Problems Relevant to Data Collection" (see below text table) and that this is included in the report of the AWG.
- In completing the form, the contact person should, where possible, indicate the course of action that they feel is required in order to address the issues identified.
- Provide feedback from PGCCDBS and RCMs to HAWG.

HAWG 2009 appointed Lotte Worsøe Clausen (DTU Aqua) as contact person for the PGCCDBS.

## Quality Assurance Framework (QAF)

The development of a Quality Assurance Framework (QAF) and associated data catalogue to strengthen link between AWGs and PGCCDBS by automating the reporting of data usage by the AWGs, reducing demands on already reduced WG time was continued. The ICES AMAWGC meeting in 2008 supported the development of a data catalogue to manage sampling meta-information so that the sampling summaries can be generated automatically. This should, at the same time, also suit the needs of STECF-SGRN when evaluating the compliance of Member States with the DCR (data collection regulation) and their National Programmes. The PGCCDBS worked on the outline of such a catalogue intersessionally and a first draft was presented at the meeting in March 2009. The implementation of the catalogue is planned to be tested by table templates which are to be supplied to selected stock coordinators.

Consequence of the new DCF sampling schemes

PGCCDBS recommended that Member States evaluate potential changes to the continuity of their stock assessment fishery data sets caused by the new DCF sampling schemes from 2009 onwards. A suitable approach could be developed around the framework for bias and precision evaluation developed by WKACCU and COST.

The evaluations should be supplied to ICES stock managers when Member States provide national assessment data for 2009, so that the assessment Working Groups can be made aware of features of the data that could explain unusual assessment model results, or to allow them to carry out sensitivity tests.

One of the new subjects within the current DCF is the obligation for MS to collect data for a list of ecosystem indicators. This list encompasses indicators like e.g. mean maximum length of the fish, distribution of fishing activities and discard rates of commercially exploited species (2008/949/EC, Appendix XIII). For the latter, PGCCDBS received a request from DG MARE, via ICES, to explain the process of building up a time series for this indicator.

PGCCDBS recommended that an additional Term of Reference should be provided to ICES stock assessment Working Groups from 2010 onwards to report on the impact of the new DCF sampling requirements on the quality and continuity of data sets used for assessments.

### 1.4.7 FRS project on factors affecting overwinter survival of larvae in the northern North Sea

The North Sea herring stock has experienced a succession of poor recruitments by the 2002, 2003, 2004, 2005, 2006 and 2007 year classes - amongst the lowest since the collapse of the stock during the mid-1970s. This has occurred despite the presence of the highest spawning-stock biomass (SSB) since 1973. In more recent years (2005-2007), the SSB has begun to decline, and fishers, managers and scientists have expressed concern for the future state of the herring stock. As a consequence, ICES advised large cuts in total allowable catches. ICES also convened the Study Group on Recruitment Variability in North Sea Planktivorous Fish (SGRECVAP), which met in January 2006 and May 2007 to consider the possible causes of the recruitment failure.

The SGRECVAP (ICES 2006a) found little or no evidence of impaired egg production by herring, or egg hatching success, from 2002 onwards. However, there was clear evidence of a large increase in mortality between the early larval phase (September/October), and the late larval phase (February the following year). The SG also noted that a similar phenomenon appeared to have occurred during the development of year classes 1988, 1989 and 1990, when the spawning biomass was also close to the recent historical maximum.

The 2006 SG noted evidence for significant changes in the transport of larvae, and in the plankton community of the North Sea, correlated with warming of the region especially since 2001. In 2007 the SG determined that hydrographic changes in the North Sea may have resulted in changes in frontal development and it is conceivable that these factors could have affected the feeding conditions of herring larvae and hence their growth and survival. The 2007 SG also determined that the timing of the changes in the plankton community was similar to those in the recruitment/survival patterns of the herring larvae (i.e., the late 1980s and around 2000).

In a search for evidence to support the empirical relationships between plankton and herring larvae survival noted by the SG, we examined the archived collections of larval herring from the Scottish MIKT surveys in the north-western North Sea which are
carried out in February each year. Specimens from surveys between 1995 and 2007 collected in a standard sampling area off the east coast of Scotland were dissected to expose the gut contents, which were identified and enumerated.

Alongside the gut contents analysis, results from a bio-physical model were used to simulate year-to-year changes in the transport of herring larvae produced in September at spawning grounds around Orkney, Shetland and off the east coast of Scotland, to the sampling area in February of the following year. Using these results we were able to estimate the survival of the population of larvae in the February sampling area during their over-winter drift period.

The composition of the gut contents of larvae indicated which plankton species were being eaten immediately prior to capture in February each year. The copepods Paracalanus $s p$. and Pseudocalanus $s p$. were consistently important components of the diet, but there were no clear trends in diet composition that could be related to survival.

Plankton abundance in the water, estimated from monitoring data collected weekly off Stonehaven on the east coast of Scotland, showed marked trends in species composition, but these were not obviously related to the diet of the herring larvae.

Two intestinal parasites were found in the gut contents of the larvae. One (larvae of a tetraphyllidean cestode) was a benign parasite that is progressively accumulated with feeding but is relatively harmless to the larvae. We deduced that we can use the incidence of this parasite as an index of the cumulative feeding history of the larvae. We found that the prevalence of tetraphyllideans in the gut contents varied significantly between years and was positively correlated with feeding success. High feeding success, indicated by high prevalence of tetraphyllideans, influenced survival by offsetting the effect of the second parasite type. We conclude that variability in cumulative food intake over the life span up to February is a significant determinant of variability in survival.

The other parasite (a digenean trematode species) is also obtained from eating infected plankton, but is harmful to the larvae. In years when tetraphyllideans were rare in the gut contents, indicating poor feeding conditions, there was a negative relationship between the more harmful digenean parasite and survival. We conclude that the benefits of good feeding conditions outweigh the detrimental effects of the digenean parasite on survival. However, when feeding conditions are poor, incidence of the digenean parasite further exacerbates the impact on survival.

Our data indicate that the survival of herring larvae in the north-western North Sea has increased since 2004, in marked contrast to the well documented continued decline in survival in the North Sea as a whole. Hence, the overall decline in North Sea herring recruitment must be caused by factors which are primarily operating in the central or southern North Sea, not in the northern North Sea. Our results attribute the improved survival in the north-western North Sea to increased food consumption, presumably due to improved plankton abundance. However, we cannot use our results to say that, conversely, the impaired survival in other parts of the North Sea must be necessarily due to poor feeding conditions. We recommend a similar study on archived samples of herring larvae from the surveys in the central and southern North Sea.

### 1.5 Commercial catch data collation, sampling, and terminology

### 1.5.1 Commercial catch and sampling: data collation and handling

## Input spreadsheet and initial data processing

Since 1999 (catch data 1998), the working group members have used a spreadsheet to provide all necessary landing and sampling data. The current version used for reporting the 2008 catch data was v1.6.4. These data were then further processed with the SALLOC-application (Patterson, 1998). This program gives the needed standard outputs on sampling status and biological parameters. It also clearly documents any decisions made by the species co-ordinators for filling in missing data and raising the catch information of one nation/quarter/area with information from another data set. This allows recalculation of data in the future, or storage and analyses in other tools like InterCatch (see section 1.5.4), choosing the same (subjective) decisions currently made by the WG. Ideally, all data for the various areas should be provided on the standard spreadsheet and processed similarly, resulting in a single output file for all stocks covered by this working group. National catch data submission was due to 23 February 2009. Some nations failed to deliver their data in time, but provided them the week after. All but one nation submitted catch and sampling data via the official exchange spreadsheets, and some of them loaded data into the InterCatch database.

More information on data handling transparency, data archiving and the current methods for compiling fisheries assessment data are given in the stock annex 3 . To facilitate a long-term data storage, the group stores all relevant catch and sampling data in a separate "archive" folder on the ICES network, which is updated annually. This collection is supposed to be kept confidential as it will contain data on misreporting and unallocated catches, and will be available for WG members on request. Table 1.5.1 gives an overview of data available at present, and the source of the data. Members are encouraged to use the latest-version input spreadsheets if the reentering of catch data is required. Figure 1.5 .1 shows the separation of areas applied to data in the archive.

### 1.5.2 Sampling

## Quality of sampling for the whole area

The level of catch sampling by area is given in the table below for all herring stocks covered by HAWG (in terms of fraction of catch sampled and number of age readings per 1000 t catch). There is considerable variation between areas. Further details of the sampling quality can be found by stock in the respective sections in the report.

|  | Official | Sampled | Age | Age readings |
| :--- | :---: | :---: | :---: | :---: |
| Area | catch $(\mathrm{t})$ | catch $(\mathrm{t})$ | readings | per 1000 t |
| $\mathrm{IVa}(\mathrm{E})$ | 19462 | 16854 | 370 | 19 |
| $\mathrm{IVa}(\mathrm{W})$ | 124563 | 106188 | 5553 | 45 |
| IVb | 57362 | 31261 | 1213 | 21 |
| IVc | 2087 | 185 | 75 | 36 |
| VIId | 24422 | 18616 | 1452 | 59 |
| VIIa(N) | 4895 | 4895 | 938 | 192 |
| VIa(N) | 25216 | 9837 | 757 | 30 |
| IIIa | 38200 | 38200 | 7499 | 196 |
| Celtic Sea,VIIj | 5794 | 5794 | 3779 | 652 |
| VIaS, VIIb,c | 10237 | 10237 | 3653 | 353 |

## The EU sampling regime

HAWG has recommended for years that sampling of commercial catches should be improved for most of the stocks. The EU directive for the collection of fisheries data was implemented in 2002 for all EU member states (Commission Regulation 1639/2001). The provisions in the "data directive" define specific sampling levels per 1000 tons catch. The definitions applicable for herring and the area covered by HAWG are given below:

| Area | SAMPLING LEVEL PER 1000 t CATCH |  |  |
| :--- | :--- | :--- | :--- |
| Baltic area (IIIa (S) and IIIb-c) | 1 sample of which | 100 fish measured and | 50 aged |
| Skagerrak (IIIa (N)) | 1 sample | 100 fish measured | 100 aged |
| North Sea (IV and VIId): | 1 sample | 50 fish measured | 25 aged |
| NE Atlantic and Western Channel ICES | 1 sample | 50 fish measured | 25 aged |
| sub-areas II, V, VI, VII (excluding d) VIII, |  |  |  |
| IX, X, XII, XIV |  |  |  |

There are some exemptions to the above mentioned sampling rules if e.g. landings of a specific EU member states are less than $5 \%$ of the total EU-quota for that particular species.

The process of setting up bilateral agreements for sampling landings into foreign ports started in 2005. However, there is scope for improvement, and more of these agreements have to be negotiated, especially between EU and non-EU countries, to reach a sufficient sampling coverage of these landings. Besides of this, HAWG notes the absence of formal agreements or procedures on the exchange of data collected from samples from foreign vessels landing into different states. HAWG decided that in the absence of guidance, this should be resolved on a case by case basis, but preferred to receive guidance from PGCCDBS (see also Sec. 1.4.6).

Given the diversity of the fleets harvesting most stocks assessed by HAWG, an appropriate spread of sampling effort over the different metiers is more important to the quality of catch at age data than a sufficient overall sampling level. The WG therefore recommends that all metiers with substantial catch should be sampled (including by-catches in the industrial fisheries), that catches landed abroad should be sampled, and information on these samples should be made available to the national laboratories.

### 1.5.3 Terminology

The WG noted that the use of "age", "winter rings" and "rings" still causes confusion outside the group (and sometimes even among WG members). The WG tries to avoid this by consequently using "rings" or "ringers" instead of "age" throughout the report. It should be observed that, for autumn spawning stocks, there is a difference of one year between "age" and "rings". Further elaboration on the rationale behind this can be found in the Stock Annex 3.

### 1.5.4 Intercatch

InterCatch is a web-based system for handling fish stock assessment data. National fish stock catches are imported to InterCatch. Stock coordinators then allocate sampled catches to unsampled catches, aggregate to stock level and download the output. The InterCatch stock output can then be used as input for the assessment models." Stock coordinators used InterCatch for the first time at the 2007 Herring Assessment Working Group. Comparisons between InterCatch and conventional used systems (e.g., Salloc and spreadsheets) were carried out annually since 2007. For the most recent year, the maximum discrepancies between the systems are presented in Table 1.5.2. These are in general very small. However, at the area level, some year-classes show much larger variances. The reasons for these discrepancies have to be elucidated in more detail during an intersessional cooperation between stock-coordinators and ICES InterCatch team.

In principle, the stock coordinators found that InterCatch is a helpful tool that it has the potential to reduce errors and work load of the stock coordinators. Many improvements have been implemented. However, in terms of practical use, there are still problems. The output files from InterCatch still not do supply the WG with the same information as the conventional systems. Especially for the WBSS and NSAS there is no information on CATON and CANUM for Div. IIIa available. Consequently, InterCatch could not be used for the stocks in the Baltic Sea. InterCatch can not be used solely unless this output is produced. Thus the system is regarded as an additional back-up and archiving system, which implies an extra workload for Stockcoordinators and data submitters. This may sum to several men-weeks a year.

### 1.6 Methods Used

### 1.6.1 ICA

"Integrated Catch-at-age Analysis" (ICA: Patterson, 1998; Needle, 2000) combines a statistical separable model of fishing mortality for recent years with a conventional VPA for the more distant past. Population estimates are tuned by abundance or CPUE indices from commercial fisheries or research-vessel surveys, which may be age-structured or not as required. ICA is run using FLICA which performed the same analysis as the original version but from an FLR platform (Fisheries Library in R). FLICA was used to assess all herring stocks in HAWG with the exception of herring in VIaS and VIIb,c.

### 1.6.2 FLXSA and FLICA [recent developments of XSA and ICA in R]

The FLR (Fisheries Library in R) system (www.flr-project.org) is an attempt to implement a framework for modelling integrated fisheries systems including population dynamics, fleet behaviour, stock assessment and management objectives. The stock assessment tools in FLR can also be used on their own in the WG context. The combination of the statistical and graphical tools in R with the stock assessment aids
the exploration of input data and results. Last year's assessment, the FLICA package was adjusted to provide raw parameter estimates together with the variancecovariance matrix as standard output from ICA. With this information, the standard diagnostics of ICA were replaced with diagnostics generated within FLR. The WG decided to show results of catchability models and regression residuals as they are actually fitted. Thus, observed indices are treated as dependent variables and VPA estimates of SSB or numbers at age are considered predictor variables. This enhances the visual judgment of the quality of model fit, even though the nature of the data would suggest a reversal of predictor and dependent variables. It may be sensible to take this into account in the way the catchability models are fitted, but this would require changes in the ICA code itself. In addition, a $Q-Q$ plot to show the distribution of the log residuals as compared to a normal distribution was added to the diagnostics output.

This year new diagnostic plots were developed. In particular, plots showing the contribution to the sum of squares (SSQ) of the tuning indices and the catch by age, year and cohort provide a detailed representation of how the model is fitting the data.

### 1.6.3 MFSP, MSYPR and MFDP

Short-term predictions for the North Sea used MFSP / MSYPR that was developed three years ago in the HAWG (Skagen; WD to HAWG 2003). Other short-term predictions were carried out using the MFDP v.1a software.

### 1.6.4 STPR used for medium term projections NS herring and VIa (north) herring

Medium term projections were performed with the STPR3 software, supplemented with a version (S3S) made to ease screening over ranges of model parameter choices. The software documentation is available from ICES or as a report (Skagen, 2003). The simulation framework covers alternative scenarios for future recruitment, weight and maturity at age, assessment error, discarding and other unaccounted mortality. The harvest rules can be examined with respect to error in future assessments by assuming that the stock numbers at age, and hence the SSB on which managers make their decisions, deviates from the real state of the stock. STPR3 does this by a simple stochastic multiplier on the stock numbers as seen by decision makers. Likewise, discrepancy between the decided TAC and the catch actually taken is simulated by a common implementation multiplier. This may account for bias due to misreporting etc. Uncertainty due to measurement (i.e. sampling of the catch derivation of CPUE) estimation within the assessment process, model mis-specification and implementation error were not explicitly modelled but assigned a combined assessment error. However, varying feedback between the assessment process and the management decision making process was not included. Feedback can cause bias in the assessment to affect the management and thus the stock which in turn affects bias in the assessment.

The simple approach in STPR allows for some evaluation of the robustness of a harvest rule to such errors, but does not pretend to foresee how these errors will appear in the future. However, to be feasible, one would assume that the harvest rule still should lead to a precautionary management if these errors have an order of magnitude that has been experienced in the past. It may be noted that previous implementation error that has not been accounted for, although it will have influenced the perception of the stock in the past. Hence, implementation error should only cover
cases where it may be different from what it was in the past or already documented and explicitly included in past data.

### 1.6.5 Management simulations

F-PRESS (Fisheries Projection and Evaluation by Stochastic Simulation) is a stochastic simulation tool which can be used to develop probabilistic assessment advice or to evaluate management strategies and harvest control rules (HCRs). F-PRESS is written and runs in R and is designed to be easy to edit by end users to suit their requirements. A description of this tool can be found in the SGMAS report (ICES CM 2006, ACFM:15). Preliminary simulations for Celtic Sea herring were carried out using this tool. These simulations were used to test the medium term behaviour of the stock in a stochastic framework, assuming a range of constant catch strategies.

### 1.6.6 Separable VPA

In situations where no tuning data exist, the WG uses separable VPA, implemented in the Lowestoft Package (Darby and Flatman, 1994). This is a VPA that assumes that fishing mortality can be separated into year and age effects. HAWG screens over terminal fishing moralities in a realistic range.

### 1.7 Discarding and unaccounted mortality by Pelagic fishing Vessels

In many fisheries, fish, invertebrates and other animals are caught as by-catch and returned to the sea, a practice known as discarding. Most animals do not survive this procedure. Reasons for discarding are various and usually have economic drivers:

- Fish smaller than the minimum landing size
- Quota for this specific species has already been taken
- Fish of undesired quality, size (high-grading) or low market value
- By-caught species of no commercial value

Theoretically, the use of modern fish finding technology used to find schools of fish should result in low by-catch. However, if species mixing occurs in pelagic schools (most notable of herring and mackerel), non-target species might be discarded. Releasing unwanted catch from the net (slipping, now generally prohibited in the North Sea) or pumping unsorted catch overboard also results in discarding.

In the area considered by HAWG, four nations reported discards from their fleets in 2008. From those, Scotland, Germany and Sweden incorporated discards in the assessment data. The discard figures were raised to national landings (based on the spatial and temporal distribution of the fleet by metier), and used in the assessment of North Sea autumn spawning (see Section 2.3) and VIaN (see Section 5.1.3) herring. For the Netherlands, the estimates of herring discards of approximately 970 tonnes (CV=35\%) in 2008 (from a fleet whose total landings is over 300000 tonnes of fish per year in the ICES area) were not sampled at a high enough resolution to allocate the catch in individual stocks (Helmond \& van Overzee WD03; Borges et al. 2008).

In the Dutch fleet there appears to be no size selection for landed herring compared to discarded herring (Figure 1.7.1).

No other nations reported on discards of herring in the pelagic fisheries, either because they did not occur, catches were not sampled for discards or there were difficulties with raising procedures (ICES, 2007/ACFM:06). No discard estimates for the total international catch were calculated, on a basis that some of the coverage is still
not high enough. There were no other studies on unaccounted fishing mortality in herring presented to HAWG.

The inclusion of discarded catch is considered to reduce bias of the assessment and thus give more realistic values of fishing mortality and biomass. However, they might also increase the uncertainty in the assessment because the sampling level for discards is usually lower than that for landings (Dickey-Collas et al. 2007). This low sampling rate is caused by the large number of different metiers in the pelagic fishery and the difficulty of predicting behaviour of the fisheries (in terms of target species and spatial and temporal distribution). Raising discard estimates to the national landings might result in a higher bias than an area based estimate of discards from the total international fleet, if sampling is insufficient. HAWG therefore recommends that the development of methods for estimating discards should be fleet based, rather than on a national basis.

## Conclusion

HAWG has no evidence that discarding of herring is a major problem at present for the estimation of population dynamics of herring, for the conservation of the stocks covered by HAWG, or for the ecosystem as a whole.

### 1.8 Ecosystem considerations, sprat and herring

Analysis of trends in weight at age and large climatic oscillation in herring stocks
Time series of weight-at-age for North Sea, Western Scotland, Western Baltic, Irish Sea, Celtic Sea and North West of Ireland herring were collected from ICES (2008) last assessment report. A clustering (median linkage ordination analysis) analysis was performed in order to identify groups of different stocks that showed similar trends over time.

Trends in weight at age are shown in Figure 1.8.1.The Irish and Celtic Sea stocks (cluster similarity $>90 \%$ ) showed a significant decline (Pearson correlation analysis, $P$-value $<0.001$ ) in the average mean weight at age from 1970 and the trend amplified after 1985 when the values become lower than the long-term mean limit of 0.17 kg . The North West of Ireland stock and the North Sea herring showed also a significant long-term decline in mean weight at age (Pearson correlation analysis, P-value < 0.05). On the contrary, weight at age increased significantly since 1972 for the Western Scotland (Pearson correlation analysis, P-value < 0.05). Western Baltic stock exhibited smaller weight at age compared to other stocks (long-term mean around 0.10 kg ) and a stable trend over time, with a small decline occurring between 1998 and 2002. However, the time series is too short (1991-2008) to allow for any conclusion on the long term trend for this stock.

A principal components analysis (PCA) was applied to the time series of weight at age. Western Baltic herring was excluded as the time series is relatively short compared to the other stocks. The first PCA component (PCA1) explained $61 \%$ of the year variability of stock weight at age (Figure 1.8.2).

Monthly values of AMO (Atlantic Multidecadal Oscillation) index (Enfiled et al. 2001; Rayner et al. 2003) were obtained from AMO official web site (http://www.cdc.noaa.gov/data/timeseries/AMO). The time series are calculated from the Kaplan sea surface temperature (SST) dataset which is updated monthly. AMO is basically considered as an index of the long term trend of the North Atlantic temperatures.

First, a cross-correlation function was applied to the untransformed data of PCA1 and the yearly median of the AMO data set (AMOA). A significant correlation was evident at different lags with the largest correlation observed at lag 0 (1.8.3a). However, both time series exhibited a strong trend during the time period analysed (Figure 1.8.2). Therefore, both PCA1 and AMOA were detrended applying the following formula:
$\mu=x-(a+b \cdot t)$
where $x$ is the average, $t$ is the time and $a$ and $b$ are the parameters estimated directly from the data. The cross-correlation function was then re-applied on the detrended data. The results showed that after detrending, the relationship between PCA1 and AMOM was not significant (Figure 1.8.3b).

HAWG concluded that the possible link between trends in weight-at-age and climate conditions should be investigated at a finer spatial scale, using stock specific time series of monthly SST in assumed key-periods for growth and condition of herring stocks.

### 1.9 Pelagic Regional Advisory Council [Pelagic RAC]

Members of HAWG have attended meetings of the Pelagic RAC since its inauguration in 2005. HAWG considers the views of the Pelagic RAC as important, and welcomes the formation of this forum to give stakeholders a role in the advisory process. HAWG notes that the Pelagic RAC also has special representation by non-EU countries, notably from Norway.

Most relevant documents from the Pelagic RAC to ICES and the European Commission about herring assessment and management were available to HAWG.

### 1.10 Data coordination through PGCCDBS and/or the Regional Coordination Meeting (RCM)

## Assessment Working Group (AWG) recommendations

The Group reviewed AWG reports with respect to recommendations addressed to PGCCDBS and processed these for either further action/other groups (like RCM, LM). The relevant recommendations for HAWG and the PGCCDBS response is listed in the below table.

| Recommendation | addressed to |
| :--- | :--- |
| HAWG recommends that all metiers with <br> substantial catch should be sampled (including <br> by-catches in the small meshed fishery) | This is a matter for the relevant RCM's, to <br> address when considering the harmonization <br> and coordination of National Programmes. <br> ACOM members (Norway) also need to <br> consider this when setting annual sampling <br> programmes. |
| HAWG encourages further examination of the <br> observed internannual variability in maturity <br> ogive using appropriate scientific methodology | Handled by WKMOG. A Workshop on Sexual <br> Maturity Staging of Herring and Sprat <br> [WKMSHS] (Chairs: Jonna Tomkiewicz and <br> Cindy van Damme/Gerd Kraus) is planned for <br> 2009 |
| HAWG recommends a workshop on the <br> identification of clupeid fish larvae to ensure <br> data quality. This WS should especially deal with <br> possible sources of misidentification of sprat, <br> herring and other clupeid larvae | This recommendation was also referred to <br> PGIPS, and PGCCDBS was of the opinion that <br> PGIPS was the appropriate group to assess this <br> request. |

## Stock Data Problems Relevant to Data Collection

HAWG identified the following issues for further discussion by the PGCCDBS in relation to stock data problems relevant to data collection:

Request for guidance on the sampling of landings of flagged vessels landing into different states under the DCF.

HAWG found that there are no formal agreements or procedures on the exchange of data collected from samples from vessels landing abroad, e.g. how do you exchange, and at what level, the information from English catch sampled when landed in The Netherlands? Upon requesting clarification from colleagues involved in the DCF and PGCCDBS, none was forthcoming. HAWG decided that in the absence of guidance, they would suggest that this should be resolved on a case by case basis, but preferred to receive guidance from PGCCDBS. HAWG perceive that problems associated with the handling and exchange of data from sampling foreign vessels are likely to increase in frequency as the EU moves towards collecting data by fleet and metier rather than by nation.

As shown by the project EMAS and the ICES study group on market sampling methodology (ICES 1999 and 2000), individual National laboratories have different sampling procedures therefore, it would not be appropriate to raise samples obtained in the context of a particular sampling strategy by means of procedures conceived for a different strategy In the case of English flagged vessels landing in the Netherlands, it was agreed that the numbers and weights at age profiles per area and quarter obtained from raw data by the institute sampling at the port of landing would be submitted to the other institute who would be responsible for determining the best estimate of the age composition of the catch for their own flagged vessels. Crucially activities should not be replicated. Given results from the studies above, the quality of the data submitted would be considered acceptable as long as the standard protocols of the sampling institute were documented and followed. Due to language problems, these protocols may not be readily available to the receiving institutes.

| Stock | Description of <br> Data Problem | How to Be addressed? | By who |
| :--- | :--- | :--- | :--- |
| North Sea <br> herring | Guidance on the sampling of <br> landings of flagged vessels <br> landing into different states <br> under the DCF. | PGCCDBS, North Sea <br> RCM, Western RCM | PGCCDBS, North Sea <br> RCM, Western RCM |
| All stocks | Spatial data and information <br> on sampling coverage and <br> precision needs to be <br> provided and if possible used <br> in the assessment. | PGCCDBS should <br> formulate data <br> requirements | PGCCDBS and in turn <br> the DCF |

### 1.11 Stock overview

Analytical assessment could be carried out for four of these eleven stocks. Results of the assessments are presented in the subsequent sections of the report and are summarized below and in Figures 1.11.1-1.11.3.

North Sea autumn spawning herring is the largest stock assessed by this WG. It has experienced very low spawning stock biomass levels in the late 1970s when the fishery was closed for a number of years. This stock began to recover until the mid-1990s, when it appeared to decrease again rapidly. A management scheme was adopted to
halt this decline. Given this, ICES advises on the basis of the agreed EU-Norway management plan. Based on the most recent estimates of SSB and fishing mortality, ICES classifies the stock as being at risk of having reduced reproductive capacity and harvested sustainably. The SSB in autumn 2008 was estimated at 1.0 million t , and is expected to remain below $\mathrm{B}_{\mathrm{pa}}\left(1.3\right.$ million t ) in 2009. $\mathrm{F}_{2-6}$ in 2008 was estimated at 0.24 , above the target $\mathrm{F}_{2-6}$ of 0.14 . The year classes since 2002 are estimated to be among the weakest since the late 1970s. Following the agreed management plan implies catches of 164300 t for fleet A and 10400 t for fleet B in 2010 in the North Sea which is expected to lead to SSB of 1.21 million tonnes in 2011.

Western Baltic Spring Spawners (WBSS) is the only spring spawning stock assessed within this WG. It is distributed in the eastern part of the North Sea, the Skagerrak, the Kattegat and the Sub-Divisions 22, 23 and 24. Within the northern area, the stock mixes with North Sea autumn spawners. An analytical assessment demonstrates that the SBB has been stable over the last decade, although the most recent value is in the lower quartile of all observations. Fishing mortality has also been stable in the same period but is larger than any proxy of $\mathrm{F}_{\mathrm{msy}}$. Recruitment has declined consistently since 2003 and the estimated number of 0-ringers in 2008 is the lowest observed value. These poor year classes have not had a dramatic effect on the spawning stock biomass as yet, due to the comparatively large size and good growth of the 2003 year class. This year class has been the largest component of the SSB for the last three years (2006-2008) and has supported the stock during this period. However, this year class is now in decline, and will pass out of the stock in the next two years, whilst its place will be taken by the sequence of poor year classes: a continuation of the decline in SSB can therefore be expected in the short and medium term.

Celtic Sea herring: The herring fisheries to the south of Ireland in the Celtic Sea and in Division VIIj have been considered to exploit the same stock. For the purpose of stock assessment and management, these areas have been combined since 1982. The fishery in the eastern part of the Celtic Sea was closed in the early eighties due to poor recruitment. In recent years the assessment has been presented as indicative of trends. In 2009 HAWG put forward an analytical assessment for this stock. The assessment showed SSB being above $\mathrm{B}_{\mathrm{pa}}$ and mean $\mathrm{F}_{2-5}$ to be declining. Overall recruitment is around long term mean. The stock is recovering. However it is still very dependent on the strength of the incoming year class, which is poorly estimated.

West of Scotland herring: The stock was larger in the 1960s when the productivity of the stock was higher. The stock experienced a heavy fishery in the mid-70s following closure of the North Sea fishery. The fishery was closed before the stock collapsed. It was opened again along with the North Sea. In the mid 1990s there was substantial area misreporting of catch into this area and sampling of catch deteriorated. Area misreporting was reduced to a very low level and information on catch has improved, but in 2004 and 2005 misreporting increased again. In the absence of precautionary reference points the state of the stock cannot be evaluated. An analytical assessment shows that SSB (in 2009) is 1.8 times Blim. ICES considers that the stock is currently fluctuating at a low level and is being exploited close to $\mathrm{F}_{\text {msy }}$. Recruitment has been low since 1998.

Herring in VIa south and VIIbc are considered to consist of a mixture of autumnand winter/spring-spawning fish. The winter/spring-spawning component is distributed in the northern part of the area. The main decline in the overall stock since 1998 appears to have taken place on the autumn-spawning component, and this is particularly evident on the traditional spawning grounds in VIIb. However, there are indica-
tions that the stock is on a historically low level. The current levels of SSB and F are not precisely known, as there is no tuned assessment available for this stock. There are no sign of stock recovery in VIaS herring.

Irish Sea autumn spawning herring comprises of two spawning groups (Manx and Mourne). This stock complex experienced a very low biomass level in the late 1970s with an increase in the mid-1980s after the introduction of quotas. The stock then declined from the late 1980s to its present level. During this time period the contribution of the Mourne spawning component declined. In the past decade there have been problems in assessing the stock, partly as a consequence of the variability in spawning migrations and mixing with the Celtic Sea stock. It seems likely that the stock has been relatively stable for the last 10 years, and that fishing mortality does not appear to be increasing above the recent average. The catches have been close to TAC levels in recent years and the main fishing activity has not varied considerably. An increase in activity on the Mourne spawning area has been observed since 2006. There is some evidence of increased recruitment in the stock in most recent years.

North Sea Sprat is a short-lived species. The recruits account for a large proportion of the stock, and the fishery in a given year is very dependent on that year's incoming year class. The size of the stock has been variable with a large biomass in the early 90's followed by a sharp decline. The state of the stock is uncertain. Survey trends indicate the stock size has varied around an average level with no trend. There is no analytical assessment for this stock.

### 1.12 Structure of the report

The report below further details in each chapter the available information on the catch, fisheries and biology of the stocks and then the stock assessments, the projections, the quality of the assessments and management considerations for each stock. This information and analyses are given in chapters for each of the seven major stocks considered by HAWG. Despite this structure, it is important to realise that there are many links between the stocks and/or areas. (e.g. North Sea and herring caught in IIIa; VIaN herring and the North Sea; VIaS, VIIbc, Irish Sea and VIaN herringand Celtic Sea and Irish Sea herring).

In 2009 HAWG carried out one three assessments: western Baltic spring spawning herring. North Sea autumn spawning herring and VIaN autumn spawning herring. These were update assessments in 2009. Based on improved data availability the Celtic Sea autumn and winter spawning stock assessment was accepted by the group as an improved update of the benchmark assessment conducted in 2007. Irish Sea herring and North Sea sprat were all exploratory assessments. One stock with poor data (IIIa sprat) is described in Section 9. Two stocks, with very poor data (no catch at age sampling) and no current ongoing research are described in Section 10. These are Clyde herring and sprat in the English Channel.

### 1.13 Recommendations

Please see Annex 2.

Table 1.5.1 Available disaggregated data for the HAWG per March 2009. X: Multiple spreadsheets (usually .xls); W: WG-data national input spreadsheets (xls); D: Disfad inputs and Alloc-outputs (ascii/txt); I: Intercatch input


Table 1.5.1: Available disaggregated data for the HAWG per March 2009. continued

| Sprat in IIIa <br> (spr_kask) | 1999 | X | (W) |  |  | provided by Else Torstensen, Mar. 2000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2000 | X | (W) |  |  | provided by Else Torstensen, Mar. 2001 |
|  | 2001 | X | (W) | D |  | provided by Lotte Askgaard Worsøe, Mar. 2002 |
|  | 2002 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2003 |
|  | 2003 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2004 |
|  | 2004 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2005 |
|  | 2005 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2006 |
|  | 2006 | X | (W) | D |  | provided by Mikael van Deurs, Mar. 2007 |
|  | 2007 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2008 |
|  | 2008 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2009 |
| Sprat in the North Sea (spr_nsea) |  |  |  |  |  |  |
|  | 1999 | X | (W) |  |  | provided by Else Torstensen, Mar. 2000 |
|  | 2000 | X | (W) |  |  | provided by Else Torstensen, Mar. 2001 |
|  | 2001 | X | (W) | D |  | provided by Lotte Askgaard Worsøe, Mar. 2002 |
|  | 2002 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2003 |
|  | 2003 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2004 |
|  | 2004 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2005 |
|  | 2005 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2006 |
|  | 2006 | X | (W) | D |  | provided by Mikael van Deurs, Mar. 2007 |
|  | 2007 | X | (W) | D | I | provided by Lotte Worsøe Clausen, Mar. 2008 |
|  | 2008 | X | (W) | D | I | provided by Lotte Worsøe Clausen, Mar. 2009 |
| Sprat in VIId \& e (spr_ech) |  |  |  |  |  |  |
|  | 1999 | X | (W) |  |  | provided by Else Torstensen, Mar. 2000 |
|  | 2000 | X | (W) |  |  | provided by Else Torstensen, Mar. 2001 |
|  | 2001 | X | (W) | D |  | provided by Lotte Askgaard Worsøe, Mar. 2002 |
|  | 2002 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2003 |
|  | 2003 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2004 |
|  | 2004 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2005 |
|  | 2005 | X | (W) | D |  | provided by Lotte Worsøe Clausen, Mar. 2006 |
|  | 2006 | X | (W) | D |  | provided by Mikael van Deurs, Mar. 2007 |
|  | 2007 | X | (W) | D | I | provided by Else Torstensen, Mar. 2008 |
|  | 2008 | X | (W) | D | I | provided by Else Torstensen, Mar. 2009 |
| National Data |  |  |  |  |  |  |
| Germany: Western Balti | 1991-2000 | X |  |  |  | provided by Tomas Gröhsler, Mar. 2001 (with sampl |
| Germany: North Sea | 1995-1998 |  | W |  |  | provided by Christopher Zimmermann, Mar 2001 (w |
| Norway: Sprat | 1995-1998 |  | W |  |  | provided by Else Torstensen, Mar 2001 (without san |
| Sweden | 1990-2000 |  | W |  |  | provided by Johan Modin, Mar 2001 (without samp. |
| UK/England \& Wales | 1985-2000 | X |  |  |  | database output provided by Marinelle Basson, Mar. |
| UK/Scotland | 1990-1998 |  | W |  |  | provided by Sandy Robb/Emma Hatfield, Mar. 2002 |

Table 1.5.2 Comparison of CANUM and WECA-estimates from conventional systems and InterCatch, by stock and age-group (winter-rings).

| North Sea (47d3) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 CANUM age Sallocl |  | CANUM IC | Proportion <br> Match (\%) | 2008 WECA age salloc |  | WECA IC | Proportion <br> Match (\%) |
|  |  |  |  |  |  |  |  |
| 0 | 798284 | 798284 | 100.000 | 0 | 0.008 | 0.008 | 99.496 |
| 1 | 235022 | 235145 | 99.948 | 1 | 0.054 | 0.054 | 99.944 |
| 2 | 331772 | 331750 | 100.007 | 2 | 0.129 | 0.129 | 99.984 |
| 3 | 184771 | 184621 | 100.081 | 3 | 0.180 | 0.180 | 99.994 |
| 4 | 199069 | 198789 | 100.141 | 4 | 0.181 | 0.181 | 100.000 |
| 5 | 137529 | 137494 | 100.026 | 5 | 0.183 | 0.183 | 99.978 |
| 6 | 118349 | 118314 | 100.030 | 6 | 0.216 | 0.216 | 100.005 |
| 7 | 215542 | 215285 | 100.120 | 7 | 0.216 | 0.216 | 99.986 |
| 8 | 74339 | 74334 | 100.006 | 8 | 0.256 | 0.256 | 99.980 |
| 9+ | 42919 | 42922 | 99.994 | 9+ | 0.273 | 0.273 | 99.996 |
| West of Scotland (VIaN) |  |  |  |  |  |  |  |
| $\begin{array}{r} 2008 \\ \text { age } \end{array}$ | CANUM Sallocl | CANUM IC | Proportion Match (\%) | $\begin{array}{r} 2008 \mathrm{~V} \\ \text { age } \mathrm{s} \end{array}$ |  | WECA IC | Proportion Match (\%) |
| 1 | 0 | 0 |  | 1 | 0.000 | 0.000 |  |
| 2 | 7898 | 7899 | 99.994 | 2 | 0.171 | 0.171 | 99.877 |
| 3 | 13039 | 13040 | 99.995 | 3 | 0.206 | 0.206 | 100.136 |
| 4 | 5428 | 5427 | 100.007 | 4 | 0.231 | 0.231 | 99.853 |
| 5 | 3220 | 3220 | 99.973 | 5 | 0.231 | 0.231 | 99.896 |
| 6 | 5689 | 5689 | 100.001 | 6 | 0.249 | 0.249 | 100.084 |
| 7 | 14832 | 14832 | 99.999 | 7 | 0.253 | 0.253 | 99.937 |
| 8 | 8142 | 8143 | 99.994 | 8 | 0.284 | 0.284 | 100.011 |
| 9+ | 8969 | 8968 | 100.002 | 9+ | 0.288 | 0.288 | 100.049 |
| Irish Sea (her-nirs) |  |  |  |  |  |  |  |
| 2008 | CANUM | CANUM | Proportion | 2008 |  | WECA | Proportion |
| age | Excel | IC | Match (\%) | age |  | IC | Match (\%) |
| 1 | 8939 | 8939 | 100.000 | 1 | 0.071 | 0.071 | 99.995 |
| 2 | 18974 | 18974 | 100.000 | 2 | 0.110 | 0.110 | 99.998 |
| 3 | 7487 | 7487 | 100.000 | 3 | 0.135 | 0.135 | 99.999 |
| 4 | 2696 | 2696 | 100.000 | 4 | 0.153 | 0.153 | 99.998 |
| 5 | 2082 | 2082 | 100.000 | 5 | 0.156 | 0.156 | 100.003 |
| 6 | 1761 | 1761 | 100.000 | 6 | 0.182 | 0.182 | 99.998 |
| 7 | 328 | 328 | 100.000 | 7 | 0.196 | 0.196 | 100.003 |
| 8 | 190 | 190 | 99.999 | 8 | 0.204 | 0.204 | 100.002 |
| 9+ | 27 | 27 | 99.998 | 9+ | 0.225 | 0.225 | 100.000 |
| Her IRLW |  |  |  |  |  |  |  |
| 2008 | CANUM | CANUM | Proportion | 2008 |  | WECA | Proportion |
| age | Excel | IC | Match (\%) | age |  | IC | Match (\%) |
| 1 | 483 | 483 | 100.00\% | 1 | 0.111 | 0.111 | 100.00\% |
| 2 | 12265 | 12265 | 100.00\% | 2 | 0.148 | 0.148 | 100.00\% |
| 3 | 19661 | 19659 | 99.99\% | 3 | 0.150 | 0.150 | 100.00\% |
| 4 | 28483 | 28479 | 99.98\% | 4 | 0.166 | 0.166 | 100.00\% |
| 5 | 11110 | 11109 | 99.98\% | 5 | 0.175 | 0.175 | 100.00\% |
| 6 | 5989 | 5984 | 99.93\% | 6 | 0.185 | 0.185 | 100.00\% |
| 7 | 2738 | 2744 | 100.23\% | 7 | 0.194 | 0.194 | 99.98\% |
| 8 | 745 | 744 | 99.84\% | 8 | 0.199 | 0.199 | 100.00\% |
| 9+ | 267 | 267 | 99.85\% | 9+ | 0.241 | 0.241 | 100.07\% |
| Her IRLS (Celtic Sea) |  |  |  |  |  |  |  |
| 2008 | CANUM | CANUM | Proportion | 2008 |  | WECA | Proportion |
|  | Excel |  | Match (\%) | age |  |  | Match (\%) |
| 1 | 1288 | 1288 | 99.99\% | 1 | 0.091 | 0.091 | 100.00\% |
| 2 | 12468 | 12468 | 100.00\% | 2 | 0.120 | 0.120 | 100.00\% |
| 3 | 8144 | 8144 | 100.00\% | 3 | 0.144 | 0.144 | 100.00\% |
| 4 | 15565 | 15565 | 100.00\% | 4 | 0.156 | 0.156 | 100.00\% |
| 5 | 2328 | 2328 | 100.00\% | 5 | 0.172 | 0.172 | 100.00\% |
| 6+ | 909 | 908 | 99.93\% | 6+ | 0.193 | 0.194 | 100.17\% |



Figure 1.5.1 ICES areas as used for the assessment of herring stocks south of $62^{\circ} \mathrm{N}$. Area names in italics indicate the area separation applied to the commercial catch and sampling data kept in long term storage. "Transfer area" refers to the transfer of Western Baltic Spring Spawners caught in the North Sea to the Baltic Assessment.


Figure 1.7.1 Relative length frequency of discarded and landed herring by the sampled Dutch pelagic freezer trawler fleet between 2003 and 2008 in the ICES area.


Figure 1.8.1 Long-term trend of annual mean-weight-at-age average over the time, for the different herring stocks (1. Celtic Sea: Celtic sea + Division VIIj; 2. W. Scot: VIa (North); 3. NW. Ireland: VIaS + VIIb,c; 4. North sea: 47d3; 5. Irish sea: VIIa (North); 6. W.Baltic SS: IIIa22).


Figure 1.8.2 First component (PCA1) of a Principal Component Analysis (PCA) of the weight at age time series and median of the monthly average of AMO (Atlantic Multidecadal Oscillation).


Figure 1.8.3 Cross correlation function between Principal Component Analysis (PCA) of the weight at age time series and median of the monthly average of AMO (Atlantic Multidecadal Oscillation). a) before detrending and $b$ ) after detrending.








Figure 1.11.1 WG estimates of catch (yield) of the stocks presented in HAWG 2009.


Figure 1.11.2 Spawning stock biomass estimates of the 4 herring stocks for which analytical assessments were presented in HAWG 2009. The $B_{p a}$ level (if defined) is indicated in the graphs.


Figure 1.11.3 Estimates of mean $F$ of the 4 herring stocks for which analytical assessments were presented in HAWG 2009. The $\mathrm{F}_{\mathrm{pa}}$ level (if defined) is indicated in the graphs.

### 2.1 The Fishery

### 2.1.1 ICES advice and management applicable to 2008 and 2009

According to the management plan agreed between the EU and Norway, adopted in December 1997 and amended in November 2007, efforts should be made to maintain the SSB of North Sea Autumn Spawning herring above 800000 tonnes.

The EU-Norway agreement on management of North Sea herring was updated in 2008, to adapt to the present reduced recruitment, accounting for the results by WKHMP (ICES 2008/ACOM:27). The management plan is given in stock annex 3.

The main changes were a reduced target F for juveniles and a higher trigger biomass for reducing the adult F . The revised rule specifies fishing mortalities for juveniles ( $\mathrm{F}_{0}-$ 1) and for adults ( $\mathrm{F}_{2-6}$ ) not to be exceeded, at 0.05 and 0.25 respectively, for the situation where the SSB is above 1.5 million tonnes. The current agreement has a constraint on year-to-year change of $15 \%$ in TAC, when the SSB is above 800000 t .

When the harvest rule leads to SSB below the trigger biomass ( 1.5 million tonnes), an iterative procedure is needed to find a fishing mortality and a corresponding SSB in the TAC year (see Stock Annex 3).

The final TAC adopted by the management bodies for 2008 was 201227 t for Area IV and Division VIId, whereof not more than 26661 t should be caught in Division IVc and VIId. For 2009, the total TAC was reduced by $15 \%$ to 186985 t (171 000 t for the A-Fleet), including a TAC of 23567 t for Division IVc and VIId.

The by-catch ceiling set for fleet B in the North Sea was 18806 t for 2008 and was decreased by $15 \%$ to 15985 t for 2009. As North Sea autumn spawners are also caught in Division IIIa, regulations for the fleets operating in this area have to be taken into account for the management of the WBSS stock (see Section 3). Catches of herring in the Thames estuary are not included in the TAC. For a definition of the different fleets harvesting North Sea herring see the stock annex and Section 2.7.2.

### 2.1.2 Catches in 2008

Total landings and estimated catches are given in the Table 2.1.1 for the North Sea and for each Division in Tables 2.1.2 to 2.1.5. Total working group catches per statistical rectangle and quarter are shown in Figures 2.1.1 ( $a-d$ ), the total for the year in Figure 2.1.1(e). Each nation provided most of their catch data (either official landings or working group catch) by statistical rectangle.

The catch figures in Tables 2.1.1-2.1.5 are mostly provided by WG members and may or may not reflect national catch statistics. These figures can therefore not be used for legal purposes. Denmark and Norway provided information on by-catches of herring in the industrial fishery. These are taken in the small-meshed fishery (Bfleet) under an EU quota by Denmark and are included in the A-fleet figures for Norway. Catch estimates of herring taken as by-catch by other small-mesh fisheries in the North Sea may be an underestimate. The total Working Group catch of all herring caught in the North Sea in 2008 amounted to 245000 t .

Landings of herring taken as by-catch in the Danish small-meshed fishery in the North Sea have increased by more than $20 \%$ to 8606 t as compared to last year (Ta-
ble 2.1.6). These industrial herring catches were much lower than the by-catch ceiling set by the EU (18 800 t ).

In the Norwegian industrial fishery, herring by-catch has decreased substantially in 2008 to 50 t (compared to 345 t last year).

Official catches by the human consumption fishery were 219100 t in 2008 (9 \% above the TAC). Working group catches in the human consumption fishery were 236400 t in 2008 (decreased by $38 \%$ from last year). The excess over the TAC for the human consumption fishery amounted to $35200 \mathrm{t}(17 \%)$ in the actual year.
In the southern North Sea and the Eastern Channel, the total catch of 39000 t in 2007 was slightly higher than the TAC of 37500 t . The over catch ratio increased again in 2008, when the catch exceed the TAC by 2900 tonnes.

The total North Sea TAC and catch estimates for the years 2000 to 2008 are shown in the table below (adapted from Table 2.1.6). Since the introduction of yearly by-catch ceilings in 1996, these ceilings have never been exceeded.

| YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TAC HC ('000 t) | 265 | 265 | 265 | 400 | 460 | 535 | 455 | 341 | 201 |
| "Official" landings HC (‘000 t) ${ }^{1}$ | 267 | 275 | 282 | 414 | 484 | 547 | 478 | 354 | 219 |
| Working Group catch HC ('000 t) | 328 | 303 | 331 | 438 | 537 | 617 | 498 | 381 | 236 |
| Excess of landings over TAC HC ('000 t) | 63 | 38 | 66 | 38 | 77 | 83 | 43 | 40 | 35 |
| By-catch ceiling (‘000 t) ${ }^{2}$ | 36 | 36 | 36 | 52 | 38 | 50 | 42 | 32 | 19 |
| Reported by-catches (‘000 t) ${ }^{3}$ | 18 | 20 | 22 | 12 | 14 | 22 | 12 | 7 | 9 |
| Working Group catch North Sea (‘000 t) | 346 | 323 | 353 | 450 | 550 | 639 | 511 | 388 | 245 |

HC = human consumption fishery
1 Landings might be provided by WG members to HAWG before the official landings become available; they may then differ from the official catches and cannot be used for management purposes. Norwegian by-catches included in this figure.
2 by-catch ceiling for EU industrial fleets only, Norwegian by-catches included in the HC figure.
3 provided by Denmark only.

### 2.1.3 Regulations and their effects

Landings taken in the North Sea but reported from other areas such as Divisions IIa and IIIIa and from Division VIaN have decreased in 2008 compared to 2007 (from 26 000 t to 17000 t ). The estimates of the total amount of catch in excess of the TAC in the human consumption fishery (excluding within-area misreporting) was about 35000 t , which is similar to last year. Along with the reduction of TACs, which have been put into place since 2006, the proportion of catch exceeding the TAC for the human consumption has increased from $9 \%$ in 2006 up to $17 \%$ in 2008.

Following the apparent recovery of the autumn spawning North Sea herring, some regulatory measures were amended: In 2004, the total Norwegian quota and half of the EU quota for Division IIIa could be taken in the North Sea. A licence scheme introduced in 1997 by UK/Scotland to reduce misreporting between the North Sea and VIaN was relaxed. The minimal amount of target species in the EU industrial fisheries in IIIa has been reduced to $50 \%$ (for sprat, blue whiting and Norway pout). In 2009, Norway can take up to 20 \% of it's quota for Division IIIa in the North Sea.

### 2.1.4 Changes in fishing technology and fishing patterns.

There have been no major changes to fish technology and fishing patterns of the fleets that target North Sea herring.

### 2.2 Biological composition of the catch

Biological information (numbers, weight, catch (SOP) at age and relative age composition) on the catch as obtained by sampling of commercial catches is given in Tables 2.2.1 to 2.2.5. Data are given for the whole year and by quarter. Except in cases where the necessary data are missing, data are displayed separately by area for herring caught in the North Sea, Western Baltic spring spawners (only in IVaE), and the total NSAS stock, including catches in Division IIIa.

Biological information on the NSAS caught in Division IIIa was obtained using splitting procedures described in Sec. 3.2 and in the stock annex 2 . Note that splitting was only applied to the working group catch, following the correction of area misreporting.

The Tables are laid out as follows:

- Table 2.2.6: Total catches of NSAS (SOP figures), mean weights and numbers-at-age by fleet
- Table 2.2.7: Data on catch numbers-at-age and SOP catches for the period 19932008 (herring caught in the North Sea)
- Table 2.2.8: WBSS taken in the North Sea (see below)
- Table 2.2.9: NSAS caught in Division IIIa
- Table 2.2.10: Total numbers of NSAS
- Table 2.2.11: Mean weights-at-age, separately for the different Divisions where NSAS are caught, for the period 1998 - 2008.

Note that SOP catch estimates may deviate in some instances slightly from the working group catch used for the assessment.

### 2.2.1 Catch in numbers-at-age

The total number of herring taken in the North Sea ( 2.1 billion fish) and the total number of NSAS ( 2.3 billion fish) have decreased in both cases by $22 \%$, as compared to last year. 0- and 1-ringers contributed $41 \%$ of the total catch in numbers of NSAS in 2008 (Table 2.2.7). 0- and 1-ringer catch has increased by $41 \%$ as compared to 2007. Most of these herring are still taken in the B-Fleet, but the amount has somewhat increased in the A-Fleet. The majority of 0 - and 1-ringers is taken in Divisions IVb and IVc, where they account up to $70 \%$ of the total catch. Roughly $40 \%$ of the total catch in the North Sea consist of the age group 4+ winter ringers.

Western Baltic and local Division IIIa Spring-spawners (WBSS) are taken in the eastern North Sea during the summer feeding migration (see stock annex 3 and section 3.2.2). These catches are included in Table 2.1.1 and listed as IIIa type. Table 2.2.8 specifies the estimated catch numbers of WBSS caught in the North Sea, which are transferred from the North Sea assessment to the assessment of Division IIIa/Western Baltic in 1993-2008. After splitting the herring caught in the North Sea and IIIa between stocks, the total catch of North Sea Autumn spawners was 257900 tonnes.

| Area | Allocated | Unallocated | DISCARDS | Total |
| :---: | :---: | :---: | :---: | :---: |
| IVa West | 124370 | 14952 | 194 | 139516 |
| IVa East | 19461 | - | - | 19461 |
| IVb | 57332 | -904 | 30 | 56458 |
| IVc/VIId | 26509 | 3103 | - | 29612 |
|  | Total catch in the | h Sea |  | 245047 |
|  | Autumn Spawners caught in Division IIIa (SOP) |  |  | 12949 |
|  | Baltic Spring Spawners caught in the North Sea (SOP) |  |  | -124 |
|  | Blackwater Spring Spawning herring |  |  | -7 |
|  | Other Spring Spawners |  |  | 0 |
|  | Total Catch NSAS used for the assessment |  |  | 257870 |

### 2.2.2 Other Spring-spawning herring in the North Sea

Norwegian Spring-spawners and local fjord-type spring spawning herring are taken in Division IVa (East) close to the Norwegian coast under a separate TAC. These catches are not included in the Norwegian North Sea catch figures given in Tables 2.1.1 to 2.1.6, but are listed separately in the respective catch tables. Catches were 2 721 t in 2008.

Blackwater herring are caught in the Thames estuary under a separate quota and included in the catch figure for England \& Wales. Catches were only 7 t in 2008.

In recent years no larger quantities of spring spawners were reported from routine sampling of commercial catch taken in the west.

### 2.2.3 Data revisions

No data revisions were applied in this year's assessment.

### 2.2.4 Quality of catch and biological data, discards

As in previous years, some nations provided information on misreported and unallocated catches of herring in the North Sea and adjacent areas. The Working Group catch, which include estimates of all fleets (and discards and misreported or unallocated catches; see Section 1.5), was estimated to exceed the official catch by $8 \%$.

Information on discards has improved compared to 2007, but is still on a low level. The final figure for discards as used in the assessment was only 224 t , based on the raised discards for three fleets. As discards are likely to occur in all national fisheries, this figure may be an underestimate. Discard data has not been consistently available for the whole time series and was only included in the assessment when reported. Estimates of discards in the Dutch fleet are approximately 970 t (CV=35 \%) in 2008, but can't be sampled at a high enough resolution to be split between area IV, VIId, VIaN and IIa (Helmond \& van Overzee WD03; Borges et al. 2008). These are not included in the assessment (see section 1.7).

In 2008, the sampling of commercial landings covers $76 \%$ of the total catch (2007: 86 $\%$ ). However, the number of herring length and weight measured has decreased by $17 \%$ when compared to 2007, and the number of age readings has decreased by $34 \%$ (Table 2.2.12). It should be observed that "sampled catch" in Table 2.2.12 refers to the proportion of the reported catch to which sampling was applied. This figure is lim-
ited to $100 \%$ but might in fact exceed the official landings due to sampling of discards, unallocated and misreported catches.

More important than a sufficient overall sampling level is an appropriate spread of sampling effort over the different metiers (here defined as each combination of fleet/nation/area and quarter. The definition of metiers may change in 2009, when new DCF rules may be put into place). Of 93 different reported metiers, only 29 were sampled in 2008. The recommended sampling level of more than 1 sample per 1000 t catch has been met only for 21 metiers, (17 in 2007). For age readings (recommended level $>25$ fish aged per 1000 t catch) also 21 metiers appear to be sampled sufficiently (2007: 16).

On the other hand, some of the metiers yielded very little catch. In 58 metiers the catch is below 1000 t . The total catch in these metiers sums to 14419 t , so the remaining 35 metiers represents 213476 t of the official catch ( $94 \%$ ). Of these 35 metiers, 20 were sampled and 11 of them fulfil the recommended level of more than 1 sample per 1000 t catch. Also 12 metiers have more than 25 age readings per 1000 t catch and 11 metiers fulfil both criteria.

However, the catch of France, Sweden, UK/Northern Ireland and the Faroe Islands from the North Sea has not been sampled.

The WG recommends that all metiers with substantial catch should be sampled (including by-catches in the industrial fisheries), and that catches landed abroad should be sampled based on criteria provided above, and information on these samples should be made available to the national laboratories (see Section 1.5).

Table 2.1.1: Herring caught in the North Sea (Sub-area IV and Division VIId). Catch in tonnes by country, 1999 - 2008. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.

| Country | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 2 | - | - | 23 | 5 |
| Denmark ${ }^{6}$ | 61268 | 64123 | 67096 | 70825 | 78606 |
| Faroe Islands | 1977 | 915 | 1082 | 1413 | 627 |
| France | 26962 | 20952 | 24880 | 25422 | 31544 |
| Germany | 26764 | 26687 | 29779 | 27213 | 43953 |
| Netherlands | 54467 | 54341 | 51293 | 55257 | 81108 |
| Norway ${ }^{1}$ | 74071 | 72072 | 75886 | 74974 | 112481 |
| Poland |  |  | - | - | - |
| Sweden | 3241 | 3046 | 3695 | 3418 | 4781 |
| USSR/Russia | - | - | - | - | - |
| UK (England) | 11434 | 11179 | 14582 | 13757 | 18639 |
| UK (Scotland) | 29911 | 30033 | 26719 | 30926 | 40292 |
| UK (N.Ireland) | - | 996 | 1018 | 944 | 2010 |
| Unallocated landings | 43327 | 61673 | 27362 | 31552 | 31875 |
| Total landings | 333424 | 346017 | 323392 | 335724 | 445921 |
| Discards | - | - | - | 17093 | 4125 |
| Total catch | 333424 | 346017 | 323392 | 352817 | 450046 |
| Estimates of the parts of the catches which have been allocated to spring spawning stocks |  |  |  |  |  |
| IIIa type (WBSS) | 4732 | 6649 | 6449 | 6652 | 2821 |
| Thames estuary ${ }^{2}$ | 88 | 76 | 107 | 60 | 84 |
| Others ${ }^{3}$ | - | 378 | 1097 | 0 | 308 |
| Norw. Spring Spawners ${ }^{4}$ | 32106 | 25678 | 7108 | 4069 | 979 |


| Country | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Belgium | 8 | 6 | 3 | 1 | - |
| Denmark $^{6}$ | 99037 | 128380 | 102322 | 84697 | 62864 |
| Faroe Islands | 402 | 738 | 1785 | 2891 | 2014 |
| France | 34521 | 38829 | 49475 | 24909 | 30347 |
| Germany | 41858 | 46555 | 40414 | 14893 | 8095 |
| Netherlands | 96162 | 81531 | 76315 | 66393 | 23122 |
| Norway $^{1}$ | 137638 | 156802 | 135361 | 100050 | 59321 |
| Poland | - | 458 | - | - | - |
| Sweden | 5692 | 13464 | 10529 | 15448 | 13840 |
| Russia | - | 99 | - | - | - |
| UK (England) | 20855 | 25311 | 22198 | 15993 | 11717 |
| UK (Scotland) | 45331 | 73227 | 48428 | 35115 | 16021 |
| UK (N.Ireland) | 2656 | 2912 | 3531 | 638 | 331 |
| Unallocated landings | 48898 | 5 | 57788 | 18764 | 26641 |
| Total landings | 533058 | 626101 | 509125 | 387669 | 244823 |
| Discards | 17059 | 12824 | 1492 | 93 | 224 |
| Total catch | 550117 | 638925 | 510617 | 387762 | 245047 |


| Estimates of the parts of the catches which have been allocated to spring spawning stocks |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
| IIIa type (WBSS) | 7079 | 7039 | 10954 | 1070 | 124 |
| Thames estuary $^{2}$ | 62 | 74 | 65 | 2 | 7 |
| Others $^{3}$ | 0 | 0 | 0 | 0 | 0 |
| Norw. Spring Spawners $^{4}$ | 452 | 417 | 626 | 685 | 2721 |

${ }^{1}$ Catches of Norwegian spring spawners removed (taken under a separate TAC).
${ }^{2}$ Landings from the Thames estuary area are included in the North Sea catch figure for UK (England).
3 Caught in the whole North Sea, partly included in the catch figure for The Netherlands
4 These catches (including some local fjord-type Spring Spawners) are taken by Norway under a separate quota south of $62^{\circ} \mathrm{N}$ and are not included in the Norwegian North Sea catch figure for this area.
5 may include misreported catch from VIaN and discards
6 Including any bycatches in the industrial fishery

Table 2.1.2: Herring caught in the North Sea. Catch in tonnes in Division IVa West. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.

| Country | 1999 | 2000 | 2001 | 2002 | 2003 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Denmark 1 | 15359 | 25530 | 17770 | 26422 | 48358 |
| Faroe Islands | 1977 | 205 | 192 | - | 95 |
| France | 6369 | 3210 | 8164 | 10522 | 11237 |
| Germany | 11206 | 5811 | 17753 | 15189 | 25796 |
| Netherlands | 21552 | 15117 | 17503 | 18289 | 25045 |
| Norway | 31395 | 33164 | 11653 | 10836 | 34443 |
| Sweden | 859 | 1479 | - | - | - |
| Poland |  |  | 1418 | 2397 | 2647 |
| Russia | - | - | - | - | - |
| UK (England) | 7999 | 8859 | 12283 | 10142 | 12030 |
| UK (Scotland) | 28537 | 29055 | 25105 | 30014 | 39970 |
| UK (N. Ireland) | - | 996 | 1018 | 944 | 2010 |
| Unallocated landings | 25469 | 24334 | 24725 | 14201 | 14115 |
| Misreporting from VIa North |  |  |  |  |  |
| Total Landings | 150722 | 167760 | 137584 | 138956 | 215746 |
| Discards |  |  |  | 17093 | 4125 |
| Total catch | 150722 | 167760 | 137584 | 156049 | 219871 |
|  |  |  |  |  |  |
|  |  |  |  |  |  |
| Country | 2004 | 2005 | 2006 | 2007 | 2008 |
|  | 48128 | 80990 | 60462 | 45948 | 28426 |
| Faroe Islands | - |  | 580 | 1118 | 2 |
| France | 10941 | 13474 | 18453 | 8570 | 13068 |
| Germany | 17559 | 22278 | 18605 | 4985 | 498 |
| Netherlands | 43876 | 36619 | 39209 | 42622 | 11634 |
| Norway | 36119 | 66232 | 38363 | 40279 | 40304 |
| Poland | - | 458 | - | - | - |
| Sweden | 2178 | 8261 | 4957 | 7658 | 7025 |
| Russia | - | 99 | - | - | - |
| UK (England) | 13480 | 15523 | 12031 | 11833 | 8355 |
| UK (Scotland) | 43490 | 71941 | 47368 | 35115 | 14727 |
| UK (N. Ireland) | 2656 | 2912 | 3531 | 638 | 331 |
| Unallocated landings | 28631 | 39324 | 10981 | 22215 | 14952 |
| Misreporting from VIa North |  |  |  |  |  |
| Total Landings | 247058 | 358111 | 253048 | 220981 | 139322 |
| Discards | 15794 | 10861 | 1492 | 93 | 194 |
| Total catch | 262852 | 368972 | 254540 | 221074 | 139516 |
|  |  |  |  |  |  |

${ }^{1}$ Including any by-catches in the industrial fishery
${ }^{2}$ May include misreported catch from VIaN and discards
${ }^{3}$ Including 1057 t of local spring spawners

Table 2.1.3: Herring caught in the North Sea. Catch in tonnes in Division IVa East. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.

| Country | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | 2001 | 2002 | $\mathbf{2 0 0 3}$ |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Denmark 1 | 18259 | 11300 | 18466 | 17846 | 7401 |
| Faroe Islands | - | 710 | 890 | 1365 | 359 |
| France | 115 | - | - | - | - |
| Germany | - | 29 | - | 81 | 54 |
| Netherlands | - | 38 | - | - | - |
| Norway 2 | 39977 | 38655 | 56904 | 63482 | 62306 |
| Sweden | 772 | 1177 | 517 | 568 | 1529 |
| Unallocated landings | - | 338 | 0 | 3959 | 9988 |
| Total landings | 59123 | 52247 | 76777 | 89303 | 83640 |
| Discards | - | - | - | - | - |
| Total catch | 59123 | 52247 | 76777 | 89303 | 83640 |
| Norw. Spring Spawners 4 | 32106 | 25678 | 7108 | 4069 | 979 |


| Country | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Denmark 1 | 16278 | 5761 | 8614 | 2646 | 1587 |
| Faroe Islands | - | 738 | 975 | 577 | 400 |
| France | - |  | - | - | - |
| Germany | 888 |  | 34 | - | - |
| Netherlands | - |  | - | 263 | - |
| Norway 2 | 100443 | 89925 | 90065 | 54424 | 17474 |
| UK (Scotland) | - | - | 83 | - | - |
| Sweden | 1720 | 3510 | 2857 | 640 | - |
| Unallocated landings | 0 | 0 | 0 | -96 | 3 |
| Total landings | 119329 | 99934 | 102628 | 58454 | 19461 |
| Discards | - | - | - | - | - |
| Total catch | 119329 | 99934 | $\mathbf{1 0 2 6 2 8}$ | 58454 | 19461 |
| Norw. Spring Spawners 4 | 452 | 417 | 626 | 685 | 2721 |

${ }^{1}$ Including any by-catches in the industrial fishery
${ }^{2}$ Catches of Norwegian spring spawning herring removed (taken under a separate TAC)
${ }^{3}$ Negative unallocated catches due to misreporting into other areas
${ }^{4}$ These catches (including some fjord-type spring spawners) are taken by Norway under a separate quota south of $62^{\circ} \mathrm{N}$ and are not included in the Norwegian North Sea catch figure for this area

Table 2.1.4: Herring caught in the North Sea. Catch in tonnes in Division IVb. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.

| Country | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Belgium | 1 | - | - | - | - |
| Denmark 1 | 26211 | 26825 | 30277 | 26387 | 22574 |
| Faroe Islands | - | - | - | 48 | 173 |
| France | 7634 | 10863 | 7796 | 4214 | 7918 |
| Germany | 13529 | 18818 | 8340 | 7577 | 12116 |
| Netherlands | 22343 | 26839 | 24160 | 13154 | 19115 |
| Norway | 2699 | 253 | 7329 | 656 | 15732 |
| Sweden | 1610 | 390 | 1760 | 453 | 605 |
| UK (England) | 1641 | 669 | 814 | 317 | 2632 |
| UK (Scotland) | 1374 | 978 | 1614 | 289 | 322 |
| Unallocated landings ${ }^{3}$ | -3794 | 4 | -9820 | 4 | -22885 |
| Total landings | 73248 | 75815 | 59205 | 5052 | -2401 |
| Discards 2 |  |  |  |  | 78786 |
| Total catch | 73248 | 75815 | 59205 | 57147 | $\mathbf{7 8 7 8 6}$ |


| Country | 2004 | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Belgium | - | - | - | - | - |
| Denmark 1 | 33857 | 41423 | 32277 | 35990 | 32230 |
| Faroe Islands | 402 | - | 200 | 1196 | 1612 |
| France | 10592 | 10205 | 17385 | 8421 | 9687 |
| Germany | 13823 | 14381 | 14222 | 2205 | 2415 |
| Netherlands | 23649 | 10038 | 13363 | 8550 | 904 |
| Norway | 1076 | 645 | 6933 | 5347 | 1543 |
| Sweden | 1794 | 1694 | 2715 | 7150 | 6815 |
| UK (England) | 2864 | 3869 | 4924 | 577 | 833 |
| UK (Scotland) | 1841 | 1286 | 977 | - | 1293 |
| Unallocated landings ${ }^{3}$ | 8300 | 10233 | 2364 | -203 | -904 |
| Total landings | 98198 | 93774 | 95360 | 69233 | 56428 |
| Discards 2 | 1265 | 1963 |  |  | 30 |
| Total catch | 99463 | 95737 | 95360 | 69233 | 56458 |

${ }^{1}$ Including any by-catches in the industrial fishery
${ }^{2}$ Discards partly included in unallocated landings
${ }^{3}$ Negative unallocated catches due to misreporting from other areas
${ }^{4}$ May include discards. Negative unallocated due to misreporting into other areas

Table 2.1.5: Herring caught in the North Sea. Catch in tonnes in Division IVc and VIId. These figures do not in all cases correspond to the official statistics and cannot be used for legal purposes.

| Country | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | 2003 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Belgium | 1 | 1 | - | 23 | 5 |
| Denmark | 1439 | 468 | 583 | 170 | 273 |
| France | 12844 | 6879 | 8750 | 10686 | 12389 |
| Germany | 2029 | 2029 | 3686 | 4366 | 5987 |
| Netherlands | 10572 | 12348 | 9630 | 23814 | 36948 |
| UK (England) | 1794 | 1651 | 1485 | 3298 | 3977 |
| UK (Scotland) | - | - | - | 623 | - |
| Unallocated landings | 21652 | 3 | 26822 | 3 | 25522 |
| Total landings | 50331 | 50198 | 49656 | 5336 | 8170 |
| Discards 2 |  |  |  | -318 | 67749 |
| Total catch | 50331 | 50198 | 49656 | 50318 | 67749 |
| Coastal spring spawners | 88 | 76 | 147 | 4 | 60 |
| included above 1 |  |  |  | 84 |  |


| Country | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Belgium | 8 | 6 | 3 | 1 | - |
| Denmark | 774 | 206 | 969 | 113 | 621 |
| Faroe Islands | - | - | 30 | - | - |
| France | 12988 | 15150 | 13637 | 7918 | 7592 |
| Germany | 9588 | 9896 | 7553 | 7703 | 5182 |
| Netherlands | 28637 | 34874 | 23743 | 14958 | 10584 |
| UK (England) | 4511 | 5919 | 5243 | 3583 | 2529 |
| UK (Scotland) | - | - | - | - | 1 |
| Unallocated landings | 11967 | 8231 | 5419 | 4725 | 3103 |
| Total landings | 68473 | 74282 | 56597 | 39001 | 29612 |
| Discards 2 | - | - | - | - | - |
| Total catch | 68473 | 74282 | 56597 | 39001 | 29612 |
| Coastal spring spawners | 62 | 74 | 65 | 2 | 7 |
| included above 1 |  |  |  |  |  |

${ }^{1}$ Landings from the Thames estuary area are included in the North Sea catch figure for UK (England)
${ }^{2}$ Discards partly included in unallocated landings
${ }^{3}$ May include misreported catch and discards
${ }^{4}$ Thames/Blackwater herring landings: 107 t , others included in the catch figure for The Netherlands
Table 2.1.6 ("The Wonderful Table"): HERRING in Sub-area IV, Division VIId and Division IIIa. Figures in thousand tonnes.

| Year | 1992 | 1993 |  | 1994 |  | 1995 |  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 200 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sub-Area IV and Division VIId: TAC (IV and VIId) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Recommended Divisions IVa, b | 352 | 290 | 5 | 296 | 5 | 389 | 8 | 156 | 159 | 254 | 265 | 265 | - 15 | - 15 | - 15 | - 15 | - 15 | - 15 | - 15 | -1 |  |
| Recommended Divisions IVc, VIId | 54 | 50 |  | 50 |  | 50 |  | - 11 | - 11 | - 11 | - 11 | - 11 | - 11 | 11 | -11 | - 11 | 11 | - 11 | - 11 | - 11 |  |
| Expected catch of spring spawners | 10 | 8 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Agreed Divisions IVa,b 1 | 380 | 380 |  | 390 |  | 390 |  | 263;131 10 | 134 | 229 | 240 | 240 | 240 | 223 | 340.5 | 393.9 | 460.7 | 404.7 | 303.5 | 174.6 | 147. |
| Agreed Div. IVc, VIId | 50 | 50 |  | 50 |  | 50 |  | 50; 2510 | 25 | 25 | 25 | 25 | 25 | 42.7 | 59.5 | 66.1 | 74.3 | 50.0 | 37.5 | 26.7 | 23. |
| Bycatch ceiling in the small mesh fishery |  |  |  |  |  |  |  |  | 24 | 22 | 30 | 36 | 36 | 36 | 52.0 | 38.0 | 50.0 | 42.5 | 31.9 | 18.8 | 16.0 |
| CATCH (IV and VIId) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| National landings Divisions IVa,b 2 | 481 | 463 |  | 421 |  | 465 |  | 183 | 149 | 245 | 261 | 261 | 272 | 261 | 354.5 | 427.7 | 502.3 | 439.2 | 326.8 | 201.2 |  |
| Unallocated landings Divisions IVa,b | 14 | -1 |  | 6 |  | -15 |  | -5 | 36 | 44 | 22 | 35 | 2 | 24 | 23.7 | 36.9 | 49.6 | 13.3 | 21.9 | 14.0 |  |
| Discard/slipping Divisions IVa,b 3 | 3 | 1 |  | 1 |  | - |  | - | - | - | - | - | - | 17 | 4.1 | 17.1 | 12.8 | 1.5 | 0.1 | 0.2 |  |
| Total catch Divisions IVa, b 4 | 498 | 463 |  | 428 |  | 450 |  | 178 | 185 | 289 | 283 | 296 | 273 | 303 | 382.3 | 481.6 | 564.6 | 454.0 | 348.8 | 215.4 |  |
| National landings Divisions IVc, VIId 3 | 37 | 32 |  | 42 |  | 45 |  | 24 | 26 | 23 | 29 | 23 | 24 | 43 | 59.5 | 56.5 | 66.1 | 51.2 | 34.3 | 26.5 |  |
| Unallocated landings Divisions IVc,VIId | 35 | 43 |  | 30 |  | 22 |  | 31 | 27 | 27 | 22 | 27 | 26 | 7 | 8.2 | 12.0 | 8.2 | 5.4 | 4.7 | 3.1 |  |
| Discard/slipping Divisions IVc, VIId 3 | 2 | 2 |  | 2 |  | - |  | - | - | - | - | - | - | 0 | - | - | - | - | - | - |  |
| Total catch Divisions IVc, VIId | 74 | 77 |  | 74 |  | 67 |  | 55 | 53 | 49 | 50 | 50 | 50 | 50 | 67.7 | 68.5 | 74.3 | 56.6 | 39.0 | 29.6 |  |
| Total catch IV and VIId as used by ICES 4 | 572 | 540 |  | 498 |  | 516 |  | 233 | 238 | 338 | 333 | 346 | 323 | 353 | 450.0 | 550.1 | 638.9 | 510.6 | 387.8 | 245.0 |  |
| CATCH BY FLEET/STOCK (IV and VIId) 7 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| North Sea autumn spawners directed fisheries (Fleet A) | 441 | 438 |  | 447 |  | 439 |  | 195 | 225 | 316 | 313 | 322 | 296 | 323 | 434.9 | 529.5 | 610.0 | 487.1 | 379.6 | 236.3 |  |
| North Sea autumn spawners industrial (Fleet B) | 124 | 101 |  | 38 |  | 67 |  | 38 | 13 | 14 | 15 | 18 | 20 | 22 | 12.3 | 13.6 | 21.8 | 11.9 | 7.1 | 8.6 |  |
| North Sea autumn spawners in IV and VIId total | 564 | 539 |  | 485 |  | 506 |  | 233 | 237 | 330 | 329 | 339 | 317 | 346 | 447.2 | 543.0 | 631.9 | 499.0 | 386.7 | 244.9 |  |
| Baltic-IIIa-type spring spawners in IV | 8 | 9 |  | 13 |  | 10 |  | 1 | 1 | 8 | 5 | 7 | 6 | 7 | 2.8 | 7.1 | 7.0 | 11.0 | 1.1 | 0.1 |  |
| Coastal-type spring spawners | 0.2 | 0.2 |  | 0.2 |  | 0.2 |  | 0.2 | 0.2 | 0.1 | 0.1 | 0.1 | 1.2 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.0 | 0.0 |  |
| Norw. Spring Spawners caught under a separate quota in IV 14 | 5 | 9 |  | , |  | 10 |  | 30 | 55 | 29 | 32 | 26 | 7 | 4 | 1.0 | 0.5 | 0.4 | 0.6 | 0.7 | 2.7 |  |
| Division IIIa: TAC (IIIa) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Predicted catch of autumn spawners | 153 | 102 |  | 77 |  | 98 |  | 48 | 35 | 58 | 43 | 53 | 15 | 5 | 15 | 15 | 15 | 15 | 15 | 15 |  |
| Recommended spring spawners | 90 | 93-113 |  |  |  |  |  | - 9 | - 12 | - 12 | - 12 | - 12 | - 12 | - 12 | - 12 | - 15 | - 15 | - 15 | - 15 | 15 |  |
| Recommended mixed clupeoids | 0 | 0 |  | - |  | - |  | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| Agreed herring TAC | 124 | 165 |  | 148 |  | 140 |  | 120 | 80 | 80 | 80 | 80 | 80 | 80 | 80.0 | 70.0 | 96.0 | 81.6 | 69.4 | 51.7 | 37. |
| Agreed mixed clupeoid TAC | 50 | 45 |  | 43 |  | 43 |  | 43 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Bycatch ceiling in the small mesh fishery |  |  |  |  |  |  |  |  | 20 | 17 | 19 | 21 | 21 | 21 | 21.0 | 21.0 | 24.2 | 20.5 | 15.4 | 11.5 | 8.4 |
| CATCH (IIIa) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| National landings | 227 | 214 |  | 168 |  | 157 |  | 115 | 83 | 120 | 86 | 108 | 90 | 79 | 76.0 | 61.1 | 90.8 | 88.9 | 47.3 | 38.2 |  |
| Catch as used by ICES | 227 | 214 |  | 168 |  | 140 |  | 105 | 74 | 108 | 79 | 99 | 82 | 73 | 68.1 | 52.7 | 69.6 | 51.2 | 47.4 | 38.2 |  |
| CATCH BY FLEET/STOCK (IIIa) 7 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Autumn spawners human consumption (Fleet C) | 47 | 44 |  | 42 |  | 38 |  | 24 | 21 | 59 | 28 | 36 | 34 | 17 | 24.1 | 13.4 | 22.9 | 11.6 | 16.4 | 9.2 |  |
| Autumn spawners mixed clupeoid (Fleet D) 13 | 23 | 25 |  | 12 |  | , |  | 9 | 4 | 6 | 8 | 13 | 12 | 9 | 8.4 | 10.8 | 9.0 | 3.4 | 3.4 | 3.7 |  |
| Autumn spawners other industrial landings (Fleet E) | 82 | 63 |  | 32 |  | 29 |  | 8 | 2 |  |  |  |  |  |  |  |  |  |  |  |  |
| Autumn spawners in IIIa total | 152 | 132 |  | 86 |  | 73 |  | 43 | 27 | 61 | 34 | 49 | 46 | 26 | 32.5 | 24.2 | 31.9 | 15.0 | 19.8 | 12.9 |  |
| Spring spawners human consumption (Fleet C) | 53 | 68 |  | 59 |  | 44 |  | 58 | 43 | 40 | 40 | 45 | 33 | 38 | 31.6 | 16.8 | 32.5 | 30.2 | 25.3 | 23.0 |  |
| Spring spawners mixed clupeoid (Fleet D) 13 | 2 | 1 |  | 1 |  | 2 |  | 4 | 3 | 3 | 3 | 5 | 3 | 9 | 4.0 | 11.2 | 5.1 | 5.9 | 2.3 | 2.2 |  |
| Spring spawners other industrial landings (Fleet E) | 20 | 12 |  | 24 |  | 21 |  | 2 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| Spring spawners in IIIIa total | 75 | 81 |  | 84 |  | 67 |  | 64 | 47 | 43 | 43 | 50 | 36 | 47 | 35.6 | 28.0 | 37.6 | 36.1 | 27.6 | 25.2 |  |
| North Sea autumn spawners Total as used by ICES | 716 | 671 |  | 571 |  | 579 |  | 275 | 264 | 392 | 363 | 388 | 363 | 372 | 479.7 | 567.2 | 663.8 | 514.6 | 406.5 | 257.9 |  |

${ }_{1}$ IVa,b and EC zone of IIa. 2 Provided by Working Group members. 3 Incomplete, only some countries providing discard information. 4 Includes spring spawners not included in assessment. 5 Based on $\mathrm{F}=0.3$ in directed fishery only; TAC advised for IVc, VIId subtracted. 6 130-180 for spring spawners in all areas. 7 Based on sum-of-products (number $x$ mean weight at age). 8 Status quo F catch for fleet A. 9 The catch should not exceed recent catch levels. 10 During the middle of 1996 revised to $50 \%$ of its original agreed TAC. 11 Included in IVa,b. 12 Managed in accordance with autumn spawners. 13 Fleet D and E are merged from 1999 onwards. 14 These catches
(including local fjord-type Spring Spawners) are taken by Norway under a separate quota south of $62^{\circ} \mathrm{N}$ and are not included in the Norwegian North Sea catch figure for this area. 15 See catch option tables for different fleets.

Table 2.2.1: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea and Div IIIa in 2008. Catch in numbers (millions) at age (CANUM), by quarter and division.

| WR | $\begin{array}{r} \text { Illa } \\ \text { NSAS } \end{array}$ | $\begin{array}{r} \hline \mathrm{IVa}(\mathrm{E}) \\ \text { all } \end{array}$ | IVa(E) WBBS | $\begin{gathered} \hline \mathrm{TVa(E)} \\ \text { NSAS } \\ \text { only } \\ \hline \end{gathered}$ | IVa(W) | IVb | IVc | VIId | $\begin{gathered} \hline \text { TVa \& } \\ \text { IVb } \\ \text { NSAS } \\ \hline \end{gathered}$ | IVc \& VIId | Total NSAS | Herring caught in the North Sea |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |


| Quarters: 1-4 |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 85.7 | 0.0 | 0.0 | 0.0 | 51.5 | 646.3 | 14.9 | 0.0 | 697.8 | 14.9 | 798.3 | 712.6 |
| 1 | 86.6 | 1.4 | 0.0 | 1.4 | 37.5 | 98.0 | 8.5 | 3.0 | 136.8 | 11.6 | 235.0 | 148.4 |
| 2 | 72.0 | 11.2 | 0.1 | 11.2 | 158.4 | 77.4 | 0.5 | 12.1 | 247.0 | 12.7 | 331.6 | 259.7 |
| 3 | 1.9 | 10.7 | 0.1 | 10.6 | 115.5 | 46.3 | 0.7 | 9.6 | 172.4 | 10.3 | 184.6 | 182.8 |
| 4 | 0.3 | 17.6 | 0.2 | 17.5 | 116.4 | 38.1 | 1.6 | 24.9 | 172.0 | 26.5 | 198.7 | 198.7 |
| 5 | 0.1 | 9.0 | 0.1 | 9.0 | 65.5 | 23.3 | 1.6 | 37.9 | 97.8 | 39.4 | 137.4 | 137.3 |
| 6 | 0.1 | 11.3 | 0.1 | 11.2 | 54.4 | 20.4 | 0.9 | 31.2 | 86.1 | 32.1 | 118.2 | 118.2 |
| 7 | 0.3 | 21.9 | 0.2 | 21.7 | 122.7 | 41.1 | 1.2 | 28.0 | 185.6 | 29.3 | 215.2 | 215.0 |
| 8 | 0.1 | 9.6 | 0.0 | 9.5 | 40.2 | 15.6 | 0.2 | 8.7 | 65.4 | 8.9 | 74.3 | 74.3 |
| $9+$ | 0.0 | 7.4 | 0.0 | 7.3 | 29.8 | 2.0 | 0.1 | 3.7 | 39.1 | 3.8 | 42.9 | 42.9 |
| Sum | 247.0 | 100.0 | 0.7 | 99.3 | 792.0 | 1008.6 | 30.1 | 159.1 | 1899.9 | 189.3 | 2336.2 | 2089.9 |

Quarter: 1

| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 21.9 | 1.3 | 0.0 | 1.3 | 10.6 | 1.1 | 0.0 | 0.0 | 12.9 | 0.0 | 34.8 | 12.9 |
| 2 | 52.0 | 4.0 | 0.0 | 4.0 | 33.2 | 0.9 | 0.1 | 0.0 | 38.1 | 0.1 | 90.2 | 38.2 |
| 3 | 1.1 | 1.6 | 0.0 | 1.6 | 9.7 | 0.6 | 0.4 | 1.4 | 11.9 | 1.8 | 14.8 | 13.7 |
| 4 | 0.2 | 1.6 | 0.0 | 1.6 | 4.3 | 1.0 | 1.1 | 7.1 | 6.9 | 8.2 | 15.3 | 15.2 |
| 5 | 0.0 | 1.5 | 0.0 | 1.5 | 8.3 | 1.0 | 1.0 | 5.6 | 10.8 | 6.5 | 17.3 | 17.3 |
| 6 | 0.0 | 0.6 | 0.0 | 0.6 | 0.2 | 0.4 | 0.4 | 3.0 | 1.2 | 3.5 | 4.6 | 4.6 |
| 7 | 0.3 | 1.5 | 0.0 | 1.5 | 2.6 | 0.7 | 0.8 | 7.0 | 4.8 | 7.7 | 12.8 | 12.5 |
| 8 | 0.0 | 0.1 | 0.0 | 0.1 | 0.0 | 0.0 | 0.1 | 0.9 | 0.2 | 0.9 | 1.2 | 1.1 |
| 9+ | 0.0 | 0.1 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.4 | 0.1 | 0.4 | 0.5 | 0.5 |
| Sum | 75.4 | 12.2 | 0.0 | 12.2 | 68.9 | 5.7 | 3.8 | 25.4 | 86.9 | 29.2 | 191.5 | 116.0 |

Quarter: 2

| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 14.9 | 0.1 | 0.0 | 0.1 | 6.3 | 16.4 | 0.0 | 0.0 | 22.8 | 0.0 | $\mathbf{3 7 . 7}$ | $\mathbf{2 2 . 8}$ |
| 2 | 14.4 | 7.0 | 0.1 | 6.9 | 79.6 | 3.1 | 0.0 | 0.0 | 89.6 | 0.0 | 104.0 | 89.7 |
| 3 | 0.3 | 7.2 | 0.1 | 7.2 | 34.2 | 1.3 | 0.0 | 0.0 | 42.7 | 0.0 | 43.0 | 42.8 |
| 4 | 0.0 | 14.6 | 0.1 | 14.5 | 53.9 | 1.5 | 0.0 | 0.1 | 69.9 | 0.1 | 70.0 | $\mathbf{7 0 . 1}$ |
| 5 | 0.0 | 6.5 | 0.1 | 6.4 | 32.4 | 1.1 | 0.0 | 0.0 | 39.9 | 0.1 | 40.0 | 40.0 |
| 6 | 0.0 | 6.4 | 0.1 | 6.4 | 12.2 | 0.6 | 0.0 | 0.0 | 19.1 | 0.0 | 19.1 | 19.2 |
| 7 | 0.0 | 14.3 | 0.1 | 14.2 | 46.7 | 1.4 | 0.0 | 0.1 | 62.3 | 0.1 | 62.4 | 62.5 |
| 8 | 0.0 | 1.6 | 0.0 | 1.6 | 4.5 | 0.3 | 0.0 | 0.0 | 6.3 | 0.0 | 6.3 | 6.3 |
| $9+$ | 0.0 | 1.2 | 0.0 | 1.2 | 3.1 | 0.1 | 0.0 | 0.0 | 4.4 | 0.0 | 4.4 | 4.4 |
| Sum | 29.7 | 58.9 | 0.5 | 58.4 | $\mathbf{2 7 2 . 9}$ | $\mathbf{2 5 . 7}$ | $\mathbf{0 . 1}$ | $\mathbf{0 . 2}$ | $\mathbf{3 5 7 . 0}$ | $\mathbf{0 . 3}$ | $\mathbf{3 8 7 . 0}$ | $\mathbf{3 5 7 . 8}$ |

Quarter: 3

| 0 | 39.3 | 0.0 | 0.0 | 0.0 | 39.5 | 559.6 | 0.0 | 0.0 | 599.2 | 0.0 | 638.5 | 599.2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 40.2 | 0.0 | 0.0 | 0.0 | 13.8 | 21.9 | 0.0 | 0.0 | 35.7 | 0.0 | 75.9 | 35.7 |
| 2 | 5.5 | 0.2 | 0.0 | 0.2 | 42.4 | 54.5 | 0.0 | 0.0 | 97.2 | 0.0 | 102.7 | 97.2 |
| 3 | 0.2 | 0.7 | 0.0 | 0.6 | 60.5 | 39.8 | 0.0 | 0.0 | 101.0 | 0.0 | 101.2 | 101.0 |
| 4 | 0.0 | 0.6 | 0.0 | 0.6 | 53.1 | 23.4 | 0.0 | 0.0 | 77.1 | 0.0 | 77.1 | 77.1 |
| 5 | 0.0 | 0.7 | 0.0 | 0.7 | 21.6 | 16.5 | 0.0 | 0.0 | 38.8 | 0.0 | 38.8 | 38.8 |
| 6 | 0.0 | 1.2 | 0.0 | 1.2 | 31.4 | 15.4 | 0.0 | 0.0 | 48.0 | 0.0 | 48.0 | 48.0 |
| 7 | 0.1 | 1.5 | 0.1 | 1.4 | 53.7 | 28.1 | 0.0 | 0.0 | 83.2 | 0.0 | 83.3 | 83.3 |
| 8 | 0.0 | 0.5 | 0.0 | 0.5 | 19.3 | 13.0 | 0.0 | 0.0 | 32.7 | 0.0 | 32.8 | 32.7 |
| 9+ | 0.0 | 0.4 | 0.0 | 0.3 | 15.9 | 0.0 | 0.0 | 0.0 | 16.3 | 0.0 | 16.3 | 16.3 |
| Sum | 85.3 | 5.7 | 0.2 | 5.5 | 351.2 | 772.3 | 0.1 | 0.0 | 1129.0 | 0.2 | 1214.5 | 1129.4 |

Quarter: 4

| 0 | 46.3 | 0.0 | 0.0 | 0.0 | 12.0 | 86.6 | 14.9 | 0.0 | 98.6 | 14.9 | 159.8 | 113.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 9.7 | 0.0 | 0.0 | 0.0 | 6.8 | 58.7 | 8.5 | 3.0 | 65.4 | 11.6 | 86.7 | 77.0 |
| 2 | 0.2 | 0.0 | 0.0 | 0.0 | 3.2 | 18.9 | 0.4 | 12.1 | 22.1 | 12.5 | 34.8 | 34.6 |
| 3 | 0.2 | 1.2 | 0.0 | 1.2 | 11.1 | 4.6 | 0.2 | 8.3 | 16.9 | 8.5 | 25.6 | 25.4 |
| 4 | 0.1 | 0.8 | 0.0 | 0.8 | 5.2 | 12.1 | 0.4 | 17.7 | 18.1 | 18.1 | 36.3 | 36.2 |
| 5 | 0.1 | 0.4 | 0.0 | 0.4 | 3.3 | 4.7 | 0.6 | 32.2 | 8.4 | 32.8 | 41.2 | 41.1 |
| 6 | 0.0 | 3.1 | 0.0 | 3.1 | 10.6 | 4.1 | 0.5 | 28.1 | 17.8 | 28.6 | 46.5 | 46.4 |
| 7 | 0.0 | 4.7 | 0.0 | 4.7 | 19.6 | 10.9 | 0.4 | 21.0 | 35.2 | 21.5 | 56.7 | 56.7 |
| 8 | 0.0 | 7.4 | 0.0 | 7.4 | 16.5 | 2.3 | 0.2 | 7.8 | 26.1 | 7.9 | 34.1 | 34.1 |
| 9+ | 0.0 | 5.7 | 0.0 | 5.7 | 10.8 | 1.9 | 0.1 | 3.3 | 18.4 | 3.3 | 21.7 | 21.7 |
| Sum | 56.6 | 23.2 | 0.0 | 23.2 | 98.9 | 204.9 | 26.1 | 133.5 | 327.1 | 159.6 | 543.3 | 486.7 |

Table 2.2.2: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea and Div IIIa in 2008. Mean weight-at-age (kg) in the catch (WECA), by quarter and division.

| WR | $\begin{array}{r} \text { Illa } \\ \text { NSAS } \end{array}$ | $\begin{gathered} \hline \mathrm{IVa}(\mathrm{E}) \\ \text { all } \end{gathered}$ | IVa(E) WBSS | IVa(W) | IVb | IVc | VIld | IVa \& IVb all | $\begin{gathered} \hline \text { IVc \& } \\ \text { VIId } \end{gathered}$ | Total NSAS | Herring caught in the North Sea |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarters: 1-4 |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0.016 | 0.000 | 0.000 | 0.007 | 0.007 | 0.012 | 0.000 | 0.007 | 0.012 | 0.008 | 0.007 |
| 1 | 0.058 | 0.072 | 0.000 | 0.072 | 0.042 | 0.044 | 0.107 | 0.050 | 0.060 | 0.054 | 0.051 |
| 2 | 0.087 | 0.138 | 0.142 | 0.142 | 0.142 | 0.118 | 0.120 | 0.142 | 0.120 | 0.129 | 0.141 |
| 3 | 0.109 | 0.173 | 0.165 | 0.187 | 0.172 | 0.128 | 0.159 | 0.182 | 0.157 | 0.180 | 0.180 |
| 4 | 0.139 | 0.172 | 0.166 | 0.187 | 0.185 | 0.133 | 0.157 | 0.185 | 0.156 | 0.181 | 0.181 |
| 5 | 0.168 | 0.174 | 0.168 | 0.188 | 0.191 | 0.142 | 0.174 | 0.188 | 0.173 | 0.183 | 0.183 |
| 6 | 0.176 | 0.216 | 0.192 | 0.230 | 0.222 | 0.167 | 0.188 | 0.226 | 0.188 | 0.216 | 0.216 |
| 7 | 0.204 | 0.210 | 0.199 | 0.219 | 0.228 | 0.162 | 0.193 | 0.220 | 0.192 | 0.216 | 0.216 |
| 8 | 0.198 | 0.253 | 0.203 | 0.262 | 0.265 | 0.210 | 0.215 | 0.262 | 0.215 | 0.256 | 0.256 |
| 9+ | 0.000 | 0.266 | 0.233 | 0.281 | 0.223 | 0.241 | 0.247 | 0.275 | 0.247 | 0.273 | 0.273 |

Quarter: 1

| 0 | 0.000 | 0.000 | 0.020 | 0.000 | 0.000 | 0.000 | 0.000 na | na | 0.000 | 0.000 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.034 | 0.072 | 0.024 | 0.074 | 0.029 | 0.000 | 0.000 | 0.070 na | 0.047 | 0.070 |  |
| 2 | 0.083 | 0.129 | 0.056 | 0.127 | 0.104 | 0.085 | 0.000 | 0.127 | 0.085 | 0.102 | 0.127 |
| 3 | 0.103 | 0.150 | 0.110 | 0.143 | 0.115 | 0.101 | 0.101 | 0.143 | 0.101 | 0.135 | 0.137 |
| 4 | 0.129 | 0.161 | 0.244 | 0.146 | 0.122 | 0.117 | 0.125 | 0.145 | 0.124 | 0.134 | 0.134 |
| 5 | 0.000 | 0.156 | 0.262 | 0.153 | 0.128 | 0.118 | 0.136 | 0.151 | 0.133 | 0.144 | 0.144 |
| 6 | 0.000 | 0.189 | 0.325 | 0.177 | 0.133 | 0.133 | 0.152 | 0.169 | 0.150 | 0.155 | 0.155 |
| 7 | 0.208 | 0.187 | 0.267 | 0.167 | 0.141 | 0.133 | 0.158 | 0.169 | 0.156 | 0.162 | 0.161 |
| 8 | 0.199 | 0.192 | 0.268 | 0.215 | 0.160 | 0.160 | 0.171 | 0.186 | 0.170 | 0.174 | 0.173 |
| $9+$ | 0.000 | 0.235 | 0.269 | 0.215 | 0.000 | 0.000 | 0.198 | 0.234 | 0.198 | 0.205 | 0.205 |

Quarter: 2

| Quarter. 2 |  |  | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | na | na | na |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0.000 | 0.000 | 0.000 | na |  |  |  |  |  |  |  |
| 1 | 0.031 | 0.076 | 0.067 | 0.076 | 0.016 | 0.101 | 0.000 | 0.033 | 0.101 | 0.032 | 0.033 |
| 2 | 0.091 | 0.143 | 0.144 | 0.134 | 0.143 | 0.109 | 0.000 | 0.135 | 0.109 | 0.129 | 0.135 |
| 3 | 0.109 | 0.163 | 0.163 | 0.164 | 0.163 | 0.134 | 0.101 | 0.164 | 0.122 | 0.164 | 0.164 |
| 4 | 0.121 | 0.166 | 0.166 | 0.163 | 0.165 | 0.154 | 0.125 | 0.163 | 0.135 | 0.163 | 0.163 |
| 5 | 0.132 | 0.164 | 0.163 | 0.169 | 0.171 | 0.157 | 0.136 | 0.168 | 0.143 | 0.168 | 0.168 |
| 6 | 0.141 | 0.190 | 0.189 | 0.190 | 0.205 | 0.180 | 0.152 | 0.190 | 0.161 | 0.190 | 0.190 |
| 7 | 0.154 | 0.191 | 0.191 | 0.187 | 0.194 | 0.180 | 0.158 | 0.188 | 0.164 | 0.188 | 0.188 |
| 8 | 0.164 | 0.192 | 0.192 | 0.210 | 0.227 | 0.212 | 0.171 | 0.206 | 0.186 | 0.206 | 0.206 |
| $9+$ | 0.000 | 0.235 | 0.235 | 0.195 | 0.227 | 0.000 | 0.198 | 0.207 | 0.198 | 0.207 | 0.207 |

Quarter: 3

| 0 | 0.015 | 0.000 | 0.000 | 0.006 | 0.006 | 0.000 | 0.000 | 0.006 na | 0.007 | 0.006 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.076 | 0.000 | 0.072 | 0.083 | 0.039 | 0.104 | 0.107 | 0.056 | 0.105 | 0.067 | 0.056 |
| 2 | 0.108 | 0.144 | 0.128 | 0.166 | 0.143 | 0.137 | 0.115 | 0.153 | 0.134 | 0.151 | 0.153 |
| 3 | 0.123 | 0.211 | 0.143 | 0.205 | 0.172 | 0.167 | 0.169 | 0.192 | 0.167 | 0.192 | 0.192 |
| 4 | 0.000 | 0.210 | 0.160 | 0.214 | 0.192 | 0.186 | 0.169 | 0.207 | 0.182 | 0.207 | 0.207 |
| 5 | 0.000 | 0.230 | 0.180 | 0.221 | 0.197 | 0.190 | 0.180 | 0.211 | 0.186 | 0.211 | 0.211 |
| 6 | 0.000 | 0.236 | 0.200 | 0.246 | 0.228 | 0.209 | 0.190 | 0.240 | 0.202 | 0.240 | 0.240 |
| 7 | 0.191 | 0.228 | 0.220 | 0.244 | 0.240 | 0.215 | 0.201 | 0.243 | 0.211 | 0.243 | 0.243 |
| 8 | 0.198 | 0.231 | 0.229 | 0.282 | 0.268 | 0.232 | 0.212 | 0.276 | 0.226 | 0.276 | 0.276 |
| $9+$ | 0.000 | 0.255 | 0.229 | 0.313 | 0.238 | 0.218 | 0.241 | 0.312 | 0.235 | 0.312 | 0.312 |

Quarter: 4

| 0 | 0.017 | 0.000 | 0.015 | 0.012 | 0.012 | 0.012 | 0.000 | 0.012 | 0.012 | 0.013 | 0.012 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.076 | 0.000 | 0.056 | 0.043 | 0.050 | 0.044 | 0.107 | 0.049 | 0.060 | 0.054 | 0.051 |
| 2 | 0.110 | 0.183 | 0.110 | 0.150 | 0.140 | 0.126 | 0.120 | 0.142 | 0.120 | 0.134 | 0.134 |
| 3 | 0.125 | 0.244 | 0.244 | 0.193 | 0.180 | 0.175 | 0.169 | 0.193 | 0.169 | 0.185 | 0.185 |
| 4 | 0.164 | 0.262 | 0.262 | 0.204 | 0.180 | 0.172 | 0.170 | 0.191 | 0.170 | 0.181 | 0.181 |
| 5 | 0.185 | 0.325 | 0.325 | 0.248 | 0.190 | 0.181 | 0.180 | 0.218 | 0.180 | 0.188 | 0.188 |
| 6 | 0.183 | 0.267 | 0.267 | 0.229 | 0.207 | 0.193 | 0.192 | 0.230 | 0.192 | 0.207 | 0.207 |
| 7 | 0.182 | 0.268 | 0.268 | 0.234 | 0.208 | 0.207 | 0.205 | 0.230 | 0.205 | 0.221 | 0.221 |
| 8 | 0.213 | 0.269 | 0.269 | 0.254 | 0.254 | 0.225 | 0.221 | 0.258 | 0.221 | 0.249 | 0.249 |
| $9+$ | 0.000 | 0.274 | 0.274 | 0.258 | 0.222 | 0.241 | 0.254 | 0.259 | 0.254 | 0.258 | 0.258 |

Table 2.2.3: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea in 2008. Mean length-at-age (cm) in the catch, by quarter and division.

| WR | Illa | $\mathrm{IVa}(\mathrm{E})$ | $\mathrm{IVa}(\mathrm{E})$ | $\mathrm{IVa}(\mathrm{W})$ | IVb | IVc | VIId | IVa \& |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NSAS | all | WBSS |  |  |  |  | IVb | VIId |
|  |  |  |  |  |  |  |  | all |  |

Quarters: 1-4

| 0 | n.d. | 0.0 | n.d. | 11.1 | 10.9 | 12.5 | 0.0 | 11.0 | 12.5 |
| :--- | :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | n.d. | 20.2 | n.d. | 20.5 | 17.9 | 18.9 | 22.8 | 18.7 | 19.9 |
| 2 | n.d. | 24.9 | n.d. | 25.1 | 25.1 | 24.3 | 23.9 | 25.1 | 24.0 |
| 3 | n.d. | 26.8 | n.d. | 27.3 | 26.7 | 25.9 | 26.1 | 27.1 | 26.0 |
| 4 | n.d. | 26.8 | n.d. | 27.1 | 27.3 | 26.5 | 26.3 | 27.1 | 26.3 |
| 5 | n.d. | 27.2 | n.d. | 27.3 | 27.9 | 26.6 | 27.0 | 27.4 | 27.0 |
| 6 | n.d. | 28.9 | n.d. | 29.1 | 29.1 | 27.5 | 27.8 | 29.1 | 27.8 |
| 7 | n.d. | 28.8 | n.d. | 28.6 | 29.4 | 27.7 | 28.0 | 28.8 | 28.0 |
| 8 | n.d. | 31.4 | n.d. | 30.6 | 31.0 | 29.1 | 28.9 | 30.8 | 28.9 |
| $9+$ | n.d. | 31.9 | n.d. | 31.3 | 29.0 | 29.3 | 29.8 | 31.3 | 29.8 |

Quarter: 1

| 0 | n.d. | 0.0 | n.d. | 0.0 | 0.0 | 0.0 | 0.0 | - | - |
| :--- | ---: | ---: | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | n.d. | 20.1 | n.d. | 20.3 | 15.1 | 0.0 | 0.0 | 19.9 | - |
| 2 | n.d. | 24.7 | n.d. | 24.7 | 24.0 | 23.3 | 0.0 | 24.7 | 23.3 |
| 3 | n.d. | 26.1 | n.d. | 25.9 | 25.6 | 25.4 | 24.0 | 25.9 | 24.3 |
| 4 | n.d. | 26.4 | n.d. | 25.7 | 26.2 | 26.4 | 25.7 | 25.9 | 25.8 |
| 5 | n.d. | 26.7 | n.d. | 26.6 | 26.4 | 26.3 | 26.3 | 26.6 | 26.3 |
| 6 | n.d. | 27.6 | n.d. | 26.9 | 27.2 | 27.2 | 27.2 | 27.3 | 27.2 |
| 7 | n.d. | 27.8 | n.d. | 27.4 | 27.4 | 27.3 | 27.3 | 27.5 | 27.3 |
| 8 | n.d. | 28.4 | n.d. | 28.7 | 29.3 | 29.3 | 28.1 | 28.6 | 28.2 |
| $9+$ | n.d. | 29.7 | n.d. | 28.7 | 0.0 | 0.0 | 28.7 | 29.7 | 28.7 |

Quarter: 2

| 0 | n.d. | 0.0 | n.d. | 0.0 | 0.0 | 0.0 | 0.0 | - | - |
| :--- | ---: | ---: | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | n.d. | 20.6 | n.d. | 20.6 | 13.6 | 22.8 | 0.0 | 15.6 | 22.8 |
| 2 | n.d. | 25.0 | n.d. | 24.6 | 25.3 | 23.8 | 0.0 | 24.6 | 23.8 |
| 3 | n.d. | 26.2 | n.d. | 25.9 | 26.4 | 25.9 | 24.0 | 26.0 | 25.2 |
| 4 | n.d. | 26.6 | n.d. | 26.1 | 26.5 | 27.1 | 25.7 | 26.2 | 26.2 |
| 5 | n.d. | 26.8 | n.d. | 26.5 | 27.1 | 27.2 | 26.3 | 26.6 | 26.6 |
| 6 | n.d. | 27.6 | n.d. | 27.3 | 28.3 | 28.3 | 27.2 | 27.4 | 27.5 |
| 7 | n.d. | 27.8 | n.d. | 27.1 | 28.1 | 28.4 | 27.3 | 27.3 | 27.6 |
| 8 | n.d. | 28.4 | n.d. | 28.1 | 29.6 | 30.2 | 28.1 | 28.2 | 28.9 |
| $9+$ | n.d. | 29.7 | n.d. | 27.8 | 29.4 | 0.0 | 28.7 | 28.4 | 28.7 |

Quarter: 3

| 0 | n.d. | 0.0 | n.d. | 10.7 | 10.7 | 0.0 | 0.0 | 10.7 | - |
| :--- | ---: | ---: | :--- | :--- | :--- | ---: | ---: | ---: | ---: |
| 1 | n.d. | 0.0 | n.d. | 21.4 | 17.6 | 22.8 | 22.8 | 19.1 | 22.8 |
| 2 | n.d. | 25.0 | n.d. | 26.2 | 25.1 | 24.6 | 23.9 | 25.6 | 24.5 |
| 3 | n.d. | 28.7 | n.d. | 27.8 | 26.7 | 26.3 | 26.5 | 27.4 | 26.3 |
| 4 | n.d. | 28.5 | n.d. | 28.1 | 27.7 | 27.3 | 26.6 | 28.0 | 27.1 |
| 5 | n.d. | 29.3 | n.d. | 28.4 | 28.1 | 27.4 | 27.0 | 28.3 | 27.2 |
| 6 | n.d. | 29.7 | n.d. | 29.4 | 29.4 | 28.3 | 27.6 | 29.4 | 28.0 |
| 7 | n.d. | 29.6 | n.d. | 29.3 | 29.9 | 28.5 | 28.0 | 29.5 | 28.4 |
| 8 | n.d. | 30.3 | n.d. | 30.4 | 31.1 | 29.3 | 28.4 | 30.7 | 29.0 |
| $9+$ | n.d. | 31.1 | n.d. | 31.5 | 28.8 | 28.3 | 29.3 | 31.5 | 29.0 |

Quarter: 4

| 0 | n.d. | 0.0 | n.d. | 12.5 | 12.5 | 12.5 | 0.0 | 12.5 | 12.5 |
| :--- | :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | n.d. | 0.0 | n.d. | 18.8 | 19.3 | 18.9 | 22.8 | 19.3 | 19.9 |
| 2 | n.d. | 28.6 | n.d. | 27.4 | 25.2 | 24.6 | 23.9 | 25.5 | 24.0 |
| 3 | n.d. | 30.6 | n.d. | 29.4 | 27.3 | 26.8 | 26.4 | 28.9 | 26.4 |
| 4 | n.d. | 30.8 | n.d. | 29.1 | 26.7 | 26.8 | 26.5 | 27.6 | 26.5 |
| 5 | n.d. | 31.8 | n.d. | 30.0 | 27.6 | 27.0 | 27.1 | 28.7 | 27.1 |
| 6 | n.d. | 31.6 | n.d. | 30.5 | 28.4 | 27.8 | 27.9 | 30.2 | 27.9 |
| 7 | n.d. | 32.0 | n.d. | 30.6 | 28.3 | 28.3 | 28.2 | 30.1 | 28.2 |
| 8 | n.d. | 32.2 | n.d. | 31.6 | 30.5 | 29.0 | 28.9 | 31.6 | 28.9 |
| $9+$ | n.d. | 32.5 | n.d. | 32.0 | 29.0 | 29.3 | 29.9 | 31.8 | 29.9 |

Table 2.2.4: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea and Div IIIa in 2008. Catches (tonnes) at-age (SOP figures), by quarter and division.

| WR | $\begin{array}{r} \text { IIIa } \\ \text { NSAS } \end{array}$ | $\begin{array}{r} \mathrm{IVa}(\mathrm{E}) \\ \mathrm{all} \end{array}$ | $\begin{aligned} & \text { IVa(E) } \\ & \text { WBSS } \end{aligned}$ | IVa(E) NSAS only | $\mathrm{IVa}(\mathrm{W})$ | IVb | IVc | VIId |  <br> IVb <br> NSAS | $\begin{array}{r} \text { IVc \& } \\ \text { VIId } \end{array}$ | Total NSAS | Herring caught in the North Sea |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

Quarters: 1-4

| 0 | 1.4 | 0.0 | 0.0 | 0.0 | 0.4 | 4.4 | 0.2 | 0.0 | 4.8 | 0.2 | 6.3 | 5.0 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 5.0 | 0.1 | 0.0 | 0.1 | 2.7 | 4.1 | 0.4 | 0.3 | 6.9 | 0.7 | 12.6 | 7.6 |
| 2 | 6.2 | 1.6 | 0.0 | 1.5 | 22.4 | 11.0 | 0.1 | 1.5 | 35.0 | 1.5 | 42.7 | 3.5 |
| 3 | 0.2 | 1.9 | 0.0 | 1.8 | 21.5 | 7.9 | 0.1 | 1.5 | 31.3 | 1.6 | 33.2 | 3.0 |
| 4 | 0.0 | 3.0 | 0.0 | 3.0 | 21.8 | 7.1 | 0.2 | 3.9 | 31.9 | 4.1 | 36.0 | 36.0 |
| 5 | 0.0 | 1.6 | 0.0 | 1.6 | 12.3 | 4.5 | 0.2 | 6.6 | 18.3 | 6.8 | 25.2 | 25.2 |
| 6 | 0.0 | 2.4 | 0.0 | 2.4 | 12.5 | 4.5 | 0.2 | 5.9 | 19.5 | 6.0 | 25.5 | 25.5 |
| 7 | 0.1 | 4.6 | 0.0 | 4.6 | 26.9 | 9.4 | 0.2 | 5.4 | 40.8 | 5.6 | 46.5 | 46.5 |
| 8 | 0.0 | 2.4 | 0.0 | 2.4 | 10.6 | 4.1 | 0.0 | 1.9 | 17.1 | 1.9 | 19.0 | 19.0 |
| $9+$ | 0.0 | 2.0 | 0.0 | 2.0 | 8.4 | 0.4 | 0.0 | 0.9 | 10.8 | 0.9 | 11.7 | 11.7 |
| Sum | 13.0 | 19.5 | 0.1 | 19.4 | 139.5 | 57.5 | 1.5 | 27.9 | 216.3 | 29.4 | 258.7 | 245.9 |

## Quarter: 1

| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.7 | 0.1 | 0.0 | 0.1 | 0.8 | 0.0 | 0.0 | 0.0 | 0.9 | 0.0 | 1.6 | 0.9 |
| 2 | 4.3 | 0.5 | 0.0 | 0.5 | 4.2 | 0.1 | 0.0 | 0.0 | 4.8 | 0.0 | 9.2 | 4.8 |
| 3 | 0.1 | 0.2 | 0.0 | 0.2 | 1.4 | 0.1 | 0.0 | 0.1 | 1.7 | 0.2 | 2.0 | 1.9 |
| 4 | 0.0 | 0.3 | 0.0 | 0.3 | 0.6 | 0.1 | 0.1 | 0.9 | 1.0 | 1.0 | 2.0 | 2.0 |
| 5 | 0.0 | 0.2 | 0.0 | 0.2 | 1.3 | 0.1 | 0.1 | 0.8 | 1.6 | 0.9 | 2.5 | 2.5 |
| 6 | 0.0 | 0.1 | 0.0 | 0.1 | 0.0 | 0.0 | 0.1 | 0.5 | 0.2 | 0.5 | 0.7 | 0.7 |
| 7 | 0.1 | 0.3 | 0.0 | 0.3 | 0.4 | 0.1 | 0.1 | 1.1 | 0.8 | 1.2 | 2.1 | 2.0 |
| 8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.2 | 0.2 | 0.2 |
| 9+ | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.1 | 0.1 | 0.1 |
| Sum | 5.3 | 1.8 | 0.0 | 1.8 | 8.8 | 0.6 | 0.5 | 3.6 | 11.1 | 4.0 | 20.4 | 15.2 |

Quarter: 2

| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.5 | 0.0 | 0.0 | 0.0 | 0.5 | 0.3 | 0.0 | 0.0 | 0.7 | 0.0 | 1.2 | 0.7 |
| 2 | 1.3 | 1.0 | 0.0 | 1.0 | 10.7 | 0.4 | 0.0 | 0.0 | 12.1 | 0.0 | 13.4 | 12.1 |
| 3 | 0.0 | 1.2 | 0.0 | 1.2 | 5.6 | 0.2 | 0.0 | 0.0 | 7.0 | 0.0 | 7.0 | 7.0 |
| 4 | 0.0 | 2.4 | 0.0 | 2.4 | 8.8 | 0.2 | 0.0 | 0.0 | 11.4 | 0.0 | 11.4 | 11.4 |
| 5 | 0.0 | 1.1 | 0.0 | 1.0 | 5.5 | 0.2 | 0.0 | 0.0 | 6.7 | 0.0 | 6.7 | 6.7 |
| 6 | 0.0 | 1.2 | 0.0 | 1.2 | 2.3 | 0.1 | 0.0 | 0.0 | 3.6 | 0.0 | 3.6 | 3.6 |
| 7 | 0.0 | 2.7 | 0.0 | 2.7 | 8.7 | 0.3 | 0.0 | 0.0 | 11.7 | 0.0 | 11.7 | 11.7 |
| 8 | 0.0 | 0.3 | 0.0 | 0.3 | 0.9 | 0.1 | 0.0 | 0.0 | 1.3 | 0.0 | 1.3 | 1.3 |
| 9+ | 0.0 | 0.3 | 0.0 | 0.3 | 0.6 | 0.0 | 0.0 | 0.0 | 0.9 | 0.0 | 0.9 | 0.9 |
| Sum | 1.8 | 10.2 | 0.1 | 10.1 | 43.6 | 1.8 | 0.0 | 0.0 | 55.5 | 0.0 | 57.4 | 55.6 |

Quarter: 3

| 0 | 0.6 | 0.0 | 0.0 | 0.0 | 0.2 | 3.4 | 0.0 | 0.0 | 3.6 | 0.0 | 4.2 | 3.6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 3.1 | 0.0 | 0.0 | 0.0 | 1.1 | 0.9 | 0.0 | 0.0 | 2.0 | 0.0 | 5.1 | 2.0 |
| 2 | 0.6 | 0.0 | 0.0 | 0.0 | 7.1 | 7.8 | 0.0 | 0.0 | 14.9 | 0.0 | 15.5 | 14.9 |
| 3 | 0.0 | 0.1 | 0.0 | 0.1 | 12.4 | 6.8 | 0.0 | 0.0 | 19.4 | 0.0 | 19.4 | 19.4 |
| 4 | 0.0 | 0.1 | 0.0 | 0.1 | 11.3 | 4.5 | 0.0 | 0.0 | 16.0 | 0.0 | 16.0 | 16.0 |
| 5 | 0.0 | 0.2 | 0.0 | 0.2 | 4.8 | 3.2 | 0.0 | 0.0 | 8.2 | 0.0 | 8.2 | 8.2 |
| 6 | 0.0 | 0.3 | 0.0 | 0.3 | 7.7 | 3.5 | 0.0 | 0.0 | 11.5 | 0.0 | 11.5 | 11.5 |
| 7 | 0.0 | 0.3 | 0.0 | 0.3 | 13.1 | 6.7 | 0.0 | 0.0 | 20.2 | 0.0 | 20.2 | 20.2 |
| 8 | 0.0 | 0.1 | 0.0 | 0.1 | 5.4 | 3.5 | 0.0 | 0.0 | 9.0 | 0.0 | 9.0 | 9.0 |
| 9+ | 0.0 | 0.1 | 0.0 | 0.1 | 5.0 | 0.0 | 0.0 | 0.0 | 5.1 | 0.0 | 5.1 | 5.1 |
| Sum | 4.3 | 1.3 | 0.0 | 1.2 | 68.3 | 40.4 | 0.0 | 0.0 | 109.9 | 0.0 | 114.2 | 109.9 |

Quarter: 4

| 0 | 0.8 | 0.0 | 0.0 | 0.0 | 0.1 | 1.0 | 0.2 | 0.0 | 1.2 | 0.2 | 2.1 | 1.4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.7 | 0.0 | 0.0 | 0.0 | 0.3 | 2.9 | 0.4 | 0.3 | 3.2 | 0.7 | 4.7 | 3.9 |
| 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.5 | 2.7 | 0.1 | 1.5 | 3.1 | 1.5 | 4.7 | 4.6 |
| 3 | 0.0 | 0.3 | 0.0 | 0.3 | 2.1 | 0.8 | 0.0 | 1.4 | 3.3 | 1.4 | 4.7 | 4.7 |
| 4 | 0.0 | 0.2 | 0.0 | 0.2 | 1.1 | 2.2 | 0.1 | 3.0 | 3.5 | 3.1 | 6.6 | 6.5 |
| 5 | 0.0 | 0.1 | 0.0 | 0.1 | 0.8 | 0.9 | 0.1 | 5.8 | 1.8 | 5.9 | 7.8 | 7.7 |
| 6 | 0.0 | 0.8 | 0.0 | 0.8 | 2.4 | 0.9 | 0.1 | 5.4 | 4.1 | 5.5 | 9.6 | 9.6 |
| 7 | 0.0 | 1.2 | 0.0 | 1.2 | 4.6 | 2.3 | 0.1 | 4.3 | 8.1 | 4.4 | 12.5 | 12.5 |
| 8 | 0.0 | 2.0 | 0.0 | 2.0 | 4.2 | 0.6 | 0.0 | 1.7 | 6.7 | 1.7 | 8.5 | 8.5 |
| 9+ | 0.0 | 1.6 | 0.0 | 1.6 | 2.8 | 0.4 | 0.0 | 0.8 | 4.8 | 0.8 | 5.6 | 5.6 |
| Sum | 1.6 | 6.3 | 0.0 | 6.2 | 18.9 | 14.7 | 1.0 | 24.3 | 39.8 | 25.3 | 66.7 | 65.1 |

Table 2.2.5: North Sea autumn spawning herring (NSAS), and western Baltic spring spawners (WBSS) caught in the North Sea in 2008. Percentage age composition (based on numbers, 3+ group summarised), by quarter and division.

| WR | IIIa <br> NSAS | $\mathrm{IVa}(\mathrm{E})$ all | $\begin{aligned} & \text { IVa(E) } \\ & \text { WBSS } \end{aligned}$ | IVa(E) NSAS only | $\mathrm{IVa}(\mathrm{W})$ | IVb | IVc | VIId | IVa \& IVb NSAS | IVc \& VIId | Total NSAS | Herring caught in the North Sea |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarters: 1-4 |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 34.7\% | 0.0\% | 0.0\% | 0.0\% | 6.5\% | 64.1\% | 49.3\% | 0.0\% | 36.7\% | 7.8\% | 34.2\% | 34.1\% |
| 1 | 35.1\% | 1.4\% | 0.0\% | 1.4\% | 4.7\% | 9.7\% | 28.3\% | 1.9\% | 7.2\% | 6.1\% | 10.1\% | 7.1\% |
| 2 | 29.2\% | 11.2\% | 10.1\% | 11.2\% | 20.0\% | 7.7\% | 1.8\% | 7.6\% | 13.0\% | 6.7\% | 14.2\% | 12.4\% |
| 3 | 0.8\% | 10.7\% | 12.4\% | 10.7\% | 14.6\% | 4.6\% | 2.2\% | 6.1\% | 9.1\% | 5.4\% | 7.9\% | 8.7\% |
| 4 | 0.1\% | 17.6\% | 23.9\% | 17.6\% | 14.7\% | 3.8\% | 5.2\% | 15.6\% | 9.1\% | 14.0\% | 8.5\% | 9.5\% |
| 5 | 0.1\% | 9.0\% | 11.0\% | 9.0\% | 8.3\% | 2.3\% | 5.1\% | 23.8\% | 5.1\% | 20.8\% | 5.9\% | 6.6\% |
| 6 | 0.0\% | 11.3\% | 11.6\% | 11.2\% | 6.9\% | 2.0\% | 3.1\% | 19.6\% | 4.5\% | 16.9\% | 5.1\% | 5.7\% |
| 7 | 0.1\% | 21.9\% | 26.0\% | 21.9\% | 15.5\% | 4.1\% | 4.1\% | 17.6\% | 9.8\% | 15.5\% | 9.2\% | 10.3\% |
| 8 | 0.0\% | 9.6\% | 2.9\% | 9.6\% | 5.1\% | 1.5\% | 0.7\% | 5.4\% | 3.4\% | 4.7\% | 3.2\% | 3.6\% |
| 9+ | 0.0\% | 7.4\% | 2.2\% | 7.4\% | 3.8\% | 0.2\% | 0.2\% | 2.3\% | 2.1\% | 2.0\% | 1.8\% | 2.1\% |
| Sum 3+ | 1.1\% | 87.4\% | 89.9\% | 87.4\% | 68.8\% | 18.5\% | 20.6\% | 90.5\% | 43.1\% | 79.3\% | 41.6\% | 46.4\% |

Quarter: 1

| 0 | 0.0\% | 0.0\% | - | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 29.0\% | 10.3\% | - | 10.3\% | 15.4\% | 18.8\% | 0.0\% | 0.0\% | 14.9\% | 0.0\% | 18.2\% | 11.1\% |
| 2 | 68.9\% | 32.8\% | - | 32.8\% | 48.2\% | 15.0\% | 2.7\% | 0.0\% | 43.9\% | 0.3\% | 47.1\% | 32.9\% |
| 3 | 1.5\% | 13.3\% | - | 13.3\% | 14.0\% | 10.2\% | 10.7\% | 5.4\% | 13.7\% | 6.1\% | 7.7\% | 11.8\% |
| 4 | 0.2\% | 12.9\% | - | 12.9\% | 6.3\% | 18.2\% | 29.3\% | 28.0\% | 8.0\% | 28.2\% | 8.0\% | 13.1\% |
| 5 | 0.0\% | 12.1\% | - | 12.1\% | 12.0\% | 18.3\% | 25.3\% | 22.0\% | 12.4\% | 22.4\% | 9.0\% | 14.9\% |
| 6 | 0.0\% | 4.6\% | - | 4.6\% | 0.4\% | 6.2\% | 10.7\% | 12.0\% | 1.3\% | 11.8\% | 2.4\% | 4.0\% |
| 7 | 0.3\% | 12.1\% | - | 12.1\% | 3.8\% | 12.5\% | 20.0\% | 27.4\% | 5.5\% | 26.5\% | 6.7\% | 10.8\% |
| 8 | 0.0\% | 1.1\% | - | 1.1\% | 0.0\% | 0.8\% | 1.4\% | 3.4\% | 0.2\% | 3.2\% | 0.6\% | 1.0\% |
| 9+ | 0.0\% | 0.9\% | - | 0.9\% | 0.0\% | 0.0\% | 0.0\% | 1.7\% | 0.1\% | 1.5\% | 0.3\% | 0.5\% |
| Sum 3+ | 2.1\% | 56.9\% | - | 56.9\% | 36.4\% | 66.1\% | 97.3\% | 100.0\% | 41.3\% | 99.7\% | 34.7\% | 55.9\% |

Quarter: 2

| 0 | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 50.1\% | 0.2\% | 0.0\% | 0.2\% | 2.3\% | 63.7\% | 0.3\% | 0.0\% | 6.4\% | 0.1\% | 9.7\% | 6.4\% |
| 2 | 48.6\% | 11.9\% | 12.2\% | 11.9\% | 29.2\% | 12.1\% | 13.2\% | 0.0\% | 25.1\% | 5.0\% | 26.9\% | 25.1\% |
| 3 | 1.0\% | 12.3\% | 11.3\% | 12.3\% | 12.5\% | 5.2\% | 15.0\% | 5.4\% | 12.0\% | 9.1\% | 11.1\% | 12.0\% |
| 4 | 0.0\% | 24.8\% | 23.3\% | 24.8\% | 19.8\% | 5.9\% | 22.4\% | 28.0\% | 19.6\% | 25.9\% | 18.1\% | 19.6\% |
| 5 | 0.2\% | 11.0\% | 10.9\% | 11.0\% | 11.9\% | 4.2\% | 18.5\% | 22.0\% | 11.2\% | 20.7\% | 10.3\% | 11.2\% |
| 6 | 0.0\% | 10.9\% | 11.5\% | 10.9\% | 4.5\% | 2.2\% | 9.5\% | 12.0\% | 5.3\% | 11.1\% | 4.9\% | 5.4\% |
| 7 | 0.0\% | 24.3\% | 25.8\% | 24.3\% | 17.1\% | 5.5\% | 17.5\% | 27.4\% | 17.5\% | 23.7\% | 16.1\% | 17.5\% |
| 8 | 0.0\% | 2.7\% | 2.9\% | 2.7\% | 1.6\% | 1.0\% | 3.4\% | 3.4\% | 1.8\% | 3.4\% | 1.6\% | 1.8\% |
| 9+ | 0.0\% | 2.0\% | 2.2\% | 2.0\% | 1.1\% | 0.3\% | 0.0\% | 1.7\% | 1.2\% | 1.1\% | 1.1\% | 1.2\% |
| Sum 3+ | 1.3\% | 87.9\% | 87.8\% | 87.9\% | 68.5\% | 24.2\% | 86.5\% | 100.0\% | 68.5\% | 94.8\% | 63.4\% | 68.6\% |

Quarter: 3

| 0 | 46.1\% | 0.0\% | 0.0\% | 0.0\% | 11.2\% | 72.5\% | 0.0\% | 0.0\% | 53.1\% | 0.0\% | 52.6\% | 53.1\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 47.1\% | 0.0\% | 0.0\% | 0.0\% | 3.9\% | 2.8\% | 3.1\% | 2.9\% | 3.2\% | 3.1\% | 6.2\% | 3.2\% |
| 2 | 6.4\% | 4.0\% | 5.1\% | 4.0\% | 12.1\% | 7.1\% | 23.0\% | 11.1\% | 8.6\% | 20.0\% | 8.5\% | 8.6\% |
| 3 | 0.3\% | 11.4\% | 12.2\% | 11.4\% | 17.2\% | 5.2\% | 13.9\% | 6.7\% | 8.9\% | 12.1\% | 8.3\% | 8.9\% |
| 4 | 0.0\% | 10.8\% | 25.2\% | 10.3\% | 15.1\% | 3.0\% | 18.9\% | 15.6\% | 6.8\% | 18.1\% | 6.3\% | 6.8\% |
| 5 | 0.0\% | 13.0\% | 11.8\% | 13.0\% | 6.1\% | 2.1\% | 11.6\% | 22.2\% | 3.4\% | 14.2\% | 3.2\% | 3.4\% |
| 6 | 0.0\% | 20.7\% | 12.4\% | 21.0\% | 8.9\% | 2.0\% | 10.9\% | 19.5\% | 4.2\% | 13.1\% | 4.0\% | 4.3\% |
| 7 | 0.1\% | 25.7\% | 27.9\% | 25.6\% | 15.3\% | 3.6\% | 13.8\% | 14.9\% | 7.4\% | 14.1\% | 6.9\% | 7.4\% |
| 8 | 0.0\% | 8.1\% | 3.1\% | 8.3\% | 5.5\% | 1.7\% | 4.5\% | 5.0\% | 2.9\% | 4.7\% | 2.7\% | 2.9\% |
| 9+ | 0.0\% | 6.2\% | 2.3\% | 6.4\% | 4.5\% | 0.0\% | 0.2\% | 2.1\% | 1.4\% | 0.7\% | 1.3\% | 1.4\% |
| Sum 3+ | 0.4\% | 96.0\% | 94.9\% | 96.0\% | 72.7\% | 17.6\% | 73.9\% | 86.1\% | 35.2\% | 76.9\% | 32.7\% | 35.2\% |

Quarter: 4

| 0 | 81.8\% | 0.0\% | 0.0\% | 0.0\% | 12.1\% | 42.3\% | 56.9\% | 0.0\% | 30.2\% | 9.3\% | 29.4\% | 23.3\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 17.1\% | 0.0\% | 0.0\% | 0.0\% | 6.8\% | 28.6\% | 32.7\% | 2.3\% | 20.0\% | 7.3\% | 16.0\% | 15.8\% |
| 2 | 0.3\% | 0.0\% | 0.0\% | 0.0\% | 3.2\% | 9.2\% | 1.5\% | 9.1\% | 6.7\% | 7.8\% | 6.4\% | 7.1\% |
| 3 | 0.3\% | 5.2\% | 4.0\% | 5.1\% | 11.2\% | 2.2\% | 0.8\% | 6.2\% | 5.2\% | 5.3\% | 4.7\% | 5.2\% |
| 4 | 0.1\% | 3.6\% | 1.4\% | 3.6\% | 5.2\% | 5.9\% | 1.6\% | 13.3\% | 5.5\% | 11.4\% | 6.7\% | 7.4\% |
| 5 | 0.2\% | 1.5\% | 0.0\% | 1.5\% | 3.3\% | 2.3\% | 2.1\% | 24.1\% | 2.6\% | 20.5\% | 7.6\% | 8.5\% |
| 6 | 0.1\% | 13.3\% | 0.0\% | 13.4\% | 10.7\% | 2.0\% | 1.9\% | 21.0\% | 5.5\% | 17.9\% | 8.6\% | 9.5\% |
| 7 | 0.0\% | 20.0\% | 0.0\% | 20.0\% | 19.9\% | 5.3\% | 1.7\% | 15.7\% | 10.8\% | 13.4\% | 10.4\% | 11.6\% |
| 8 | 0.0\% | 31.8\% | 0.0\% | 31.8\% | 16.6\% | 1.1\% | 0.6\% | 5.8\% | 8.0\% | 5.0\% | 6.3\% | 7.0\% |
| 9+ | 0.0\% | 24.6\% | 0.0\% | 24.6\% | 10.9\% | 0.9\% | 0.2\% | 2.5\% | 5.6\% | 2.1\% | 4.0\% | 4.5\% |
| Sum 3+ | 0.8\% | 100.0\% | 5.4\% | 100.0\% | 77.9\% | 19.9\% | 8.9\% | 88.6\% | 43.1\% | 75.6\% | 48.2\% | 53.8\% |

Table 2.2.6: Total catch of herring caught in the North Sea and Div. IIIa: North Sea autumn spawners (NSAS). Catch in numbers (millions) at mean weight-at-age (kg) by fleet, and SOP catches (' $\mathbf{\prime} 00 \mathrm{t}$ ). SOP catch might deviate from reported catch as used for the assessment.

| 2005 | Fleet A |  | Fleet B |  | Fleet C |  | Fleet D |  | TOTAL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total |  | Mean |  | Mean |  | Mean |  | Mean |  | Mean |
| Winter rings | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight |
| 0 | 0.4 | 0.119 | 918.7 | 0.011 | 11.3 | 0.027 | 85.1 | 0.015 | 1,015.6 | 0.011 |
| 1 | 42.3 | 0.088 | 365.8 | 0.033 | 174.6 | 0.065 | 132.9 | 0.032 | 715.5 | 0.044 |
| 2 | 196.3 | 0.122 | 0.0 | 0.000 | 115.9 | 0.072 | 43.3 | 0.068 | 355.4 | 0.099 |
| 3 | 469.5 | 0.155 | 0.0 | 0.000 | 12.4 | 0.106 | 3.7 | 0.105 | 485.7 | 0.153 |
| 4 | 1313.0 | 0.166 | 0.0 | 0.000 | 4.7 | 0.154 | 0.6 | 0.158 | 1,318.4 | 0.166 |
| 5 | 477.6 | 0.208 | 0.0 | 0.000 | 2.1 | 0.175 | 0.2 | 0.157 | 479.9 | 0.208 |
| 6 | 573.6 | 0.223 | 0.0 | 0.000 | 1.9 | 0.189 | 0.3 | 0.160 | 575.9 | 0.223 |
| 7 | 114.7 | 0.240 | 0.0 | 0.000 | 0.3 | 0.216 | 0.2 | 0.178 | 115.2 | 0.240 |
| 8 | 107.8 | 0.266 | 0.0 | 0.000 | 0.2 | 0.209 | 0.0 | 0.000 | 108.0 | 0.266 |
| 9+ | 39.1 | 0.265 | 0.0 | 0.000 |  |  |  |  | 39.1 | 0.265 |
| TOTAL | 3,334.2 |  | 1,284.5 |  | 323.5 |  | 266.4 |  | 5,208.7 |  |
| SOP catch |  | 611.7 |  | 21.8 |  | 22.9 |  | 9.0 |  | 665.4 |


| 2006 | Fleet A |  | Fleet B |  | Fleet C |  | Fleet D |  | TOTAL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total |  | Mean |  | Mean |  | Mean |  | Mean |  | Mean |
| Winter rings | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight |
| 0 | 7.6 | 0.065 | 835.9 | 0.010 | 6.0 | 0.020 | 29.1 | 0.013 | 878.6 | 0.010 |
| 1 | 14.3 | 0.111 | 57.8 | 0.023 | 93.3 | 0.068 | 56.8 | 0.030 | 222.2 | 0.049 |
| 2 | 334.1 | 0.127 | 20.3 | 0.044 | 42.1 | 0.081 | 8.1 | 0.069 | 404.5 | 0.117 |
| 3 | 308.2 | 0.145 | 1.0 | 0.119 | 7.3 | 0.119 | 2.9 | 0.113 | 319.4 | 0.144 |
| 4 | 471.8 | 0.172 | 3.8 | 0.153 | 2.4 | 0.141 | 0.8 | 0.137 | 478.8 | 0.172 |
| 5 | 1012.6 | 0.181 | 4.7 | 0.160 | 2.1 | 0.184 | 1.2 | 0.188 | 1,020.6 | 0.181 |
| 6 | 257.5 | 0.220 | 0.0 | 0.000 | 0.4 | 0.188 | 0.1 | 0.197 | 258.1 | 0.219 |
| 7 | 253.3 | 0.237 | 0.0 | 0.000 | 0.3 | 0.213 | 0.1 | 0.225 | 253.7 | 0.237 |
| 8 | 64.6 | 0.235 | 0.5 | 0.214 | 0.1 | 0.206 | 0.0 | 0.209 | 65.3 | 0.235 |
| 9+ | 44.7 | 0.262 | 0.0 | 0.000 |  |  |  |  | 44.7 | 0.262 |
| TOTAL | 2,768.8 |  | 924.0 |  | 154.1 |  | 99.2 |  | 3,946.0 |  |
| SOP catch |  | 497.5 |  | 11.8 |  | 11.6 |  | 3.4 |  | 524.3 |


| 2007 | Fleet A |  | Fleet B |  | Fleet C |  | Fleet D |  | TOTAL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total Winter rings | Numbers | Mean Weight | Numbers | Mean Weight | Numbers | Mean Weight | Numbers | Mean Weight | Numbers | Mean Weight |
| 0 | 20.5 | 0.008 | 532.8 | 0.011 | 14.2 | 0.048 | 53.5 | 0.021 | 621.0 | 0.012 |
| 1 | 21.0 | 0.099 | 25.2 | 0.045 | 150.3 | 0.071 | 39.0 | 0.031 | 235.6 | 0.064 |
| 2 | 142.1 | 0.149 | 0.0 | 0.000 | 59.5 | 0.075 | 17.4 | 0.059 | 219.0 | 0.121 |
| 3 | 412.8 | 0.152 | 0.0 | 0.000 | 1.9 | 0.111 | 0.2 | 0.085 | 414.8 | 0.151 |
| 4 | 284.0 | 0.164 | 0.0 | 0.000 | 0.3 | 0.123 | 0.1 | 0.130 | 284.5 | 0.163 |
| 5 | 307.4 | 0.194 | 0.0 | 0.000 | 1.4 | 0.152 | 0.1 | 0.145 | 308.9 | 0.193 |
| 6 | 628.1 | 0.190 | 0.0 | 0.000 | 0.2 | 0.179 | 0.1 | 0.191 | 628.4 | 0.190 |
| 7 | 146.8 | 0.224 | 0.0 | 0.000 | 0.6 | 0.175 | 0.0 | 0.165 | 147.5 | 0.223 |
| 8 | 132.9 | 0.235 | 0.0 | 0.000 | 0.0 | 0.144 | 0.0 | 0.216 | 132.9 | 0.235 |
| 9+ | 23.2 | 0.252 | 0.0 | 0.000 | 0.0 | 0.000 | 0.0 | 0.000 | 23.2 | 0.252 |
| TOTAL | 2,118.9 |  | 558.1 |  | 228.4 |  | 110.4 |  | 3,015.8 |  |
| SOP catch |  | 381.1 |  | 6.9 |  | 16.4 |  | 3.4 |  | 407.8 |


| 2008 | Fleet A |  | Fleet B |  | Fleet C |  | Fleet D |  | TOTAL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total |  | Mean |  | Mean |  | Mean |  | Mean |  | Mean |
| Winter rings | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight | Numbers | Weight |
| 0 | 66.3 | 0.010 | 646.3 | 0.007 | 4.3 | 0.036 | 81.3 | 0.015 | 798.3 | 0.008 |
| 1 | 78.4 | 0.061 | 70.1 | 0.040 | 59.2 | 0.071 | 27.4 | 0.029 | 235.0 | 0.053 |
| 2 | 259.7 | 0.141 | 0.0 | 0.000 | 52.6 | 0.087 | 19.4 | 0.085 | 331.7 | 0.129 |
| 3 | 182.8 | 0.180 | 0.0 | 0.000 | 1.7 | 0.109 | 0.2 | 0.110 | 184.7 | 0.180 |
| 4 | 198.7 | 0.181 | 0.0 | 0.000 | 0.2 | 0.139 | 0.0 | 0.133 | 198.9 | 0.181 |
| 5 | 137.3 | 0.183 | 0.0 | 0.000 | 0.1 | 0.168 | 0.0 | 0.187 | 137.5 | 0.183 |
| 6 | 118.2 | 0.216 | 0.0 | 0.000 | 0.1 | 0.175 | 0.0 | 0.161 | 118.3 | 0.216 |
| 7 | 215.0 | 0.216 | 0.0 | 0.000 | 0.3 | 0.203 | 0.0 | 0.184 | 215.4 | 0.216 |
| 8 | 74.3 | 0.256 | 0.0 | 0.000 | 0.1 | 0.199 | 0.0 | 0.159 | 74.3 | 0.256 |
| 9+ | 42.9 | 0.273 | 0.0 | 0.000 | 0.0 | 0.000 | 0.0 | 0.000 | 42.9 | 0.273 |
| TOTAL | 1,373.6 |  | 716.4 |  | 118.6 |  | 128.3 |  | 2,336.9 |  |
| SOP catch |  | 238.7 |  | 7.1 |  | 9.2 |  | 3.7 |  | 258.8 |

Table 2.2.7: Catch at age (numbers in millions) of herring caught in the North Sea, 1993-2008. SG Rednose's revisions for 1995-2001 are included.

| Year/rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | $9+$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1993 | 7254 | 1385 | 792 | 614 | 315 | 222 | 230 | 191 | 88 | 42 |
| 1994 | 3834 | 497 | 1438 | 504 | 355 | 117 | 98 | 78 | 71 | 46 |
| 1995 | 6294 | 484 | 1319 | 818 | 244 | 122 | 57 | 43 | 69 | 29 |
| 1996 | 1795 | 645 | 488 | 516 | 170 | 57 | 22 | 9 | 17 | 4 |
| 1997 | 364 | 174 | 565 | 428 | 285 | 109 | 31 | 12 | 19 | 6 |
| 1998 | 208 | 254 | 1084 | 525 | 267 | 179 | 89 | 14 | 17 | 4 |
| 1999 | 968 | 73 | 487 | 1034 | 289 | 134 | 70 | 28 | 10 | 2 |
| 2000 | 873 | 194 | 516 | 453 | 636 | 212 | 82 | 36 | 15 | 3 |
| 2001 | 1025 | 58 | 678 | 473 | 279 | 319 | 92 | 39 | 18 | 2 |
| 2002 | 319 | 490 | 513 | 913 | 294 | 136 | 164 | 47 | 34 | 7 |
| 2003 | 347 | 172 | 1022 | 507 | 809 | 244 | 106 | 121 | 37 | 303 |
| 2004 | 627 | 136 | 274 | 1333 | 517 | 721 | 170 | 100 | 70 | 22 |
| 2005 | 919 | 408 | 203 | 487 | 1326 | 480 | 577 | 116 | 108 | 39 |
| 2006 | 844 | 72 | 354 | 309 | 475 | 1017 | 257 | 252 | 65 | 44 |
| 2007 | 553 | 46 | 142 | 413 | 284 | 307 | 628 | 147 | 133 | 23 |
| 2008 | 713 | 148 | 260 | 183 | 199 | 137 | 118 | 215 | 74 | 43 |

Table 2.2.8: Catch at age (numbers in millions) of Baltic Spring spawning Herring taken in the North Sea, and transfered to the assessment of the spring spawning stock in IIIa, 1993-2008.

| Year/rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.0 | 0.0 | 4.2 | 10.8 | 12.3 | 8.4 | 5.9 | 4.7 | 1.7 | 1.0 | 49.0 |
| 1994 | 0.0 | 0.0 | 8.8 | 28.2 | 16.3 | 11.0 | 8.6 | 3.4 | 3.2 | 0.7 | 80.2 |
| 1995 | 0.0 | 0.0 | 22.4 | 11.0 | 14.9 | 4.0 | 2.9 | 1.9 | 0.7 | 0.0 | 57.8 |
| 1996 | 0.0 | 0.0 | 0.0 | 2.8 | 0.8 | 0.4 | 0.1 | 0.1 | 0.3 | 0.0 | 4.5 |
| 1997 | 0.0 | 0.0 | 2.2 | 1.3 | 1.5 | 0.4 | 0.2 | 0.1 | 0.2 | 0.0 | 5.9 |
| 1998 | 0.0 | 5.1 | 9.5 | 12.0 | 10.1 | 6.0 | 3.0 | 0.4 | 0.9 | 0.0 | 47.0 |
| 1999 | 0.0 | 0.0 | 3.3 | 14.3 | 5.6 | 3.6 | 1.4 | 0.6 | 0.4 | 0.0 | 29.3 |
| 2000 | 0.0 | 0.0 | 8.2 | 9.8 | 10.2 | 5.7 | 2.5 | 0.6 | 0.7 | 0.1 | 37.6 |
| 2001 | 0.0 | 0.0 | 11.3 | 10.2 | 6.1 | 7.2 | 2.7 | 1.6 | 0.4 | 0.0 | 39.9 |
| 2002 | 0.0 | 0.0 | 7.6 | 14.8 | 10.6 | 3.3 | 2.9 | 1.0 | 0.5 | 0.1 | 40.8 |
| 2003 | 0.0 | 0.0 | 0.0 | 3.1 | 6.0 | 3.5 | 1.2 | 1.3 | 0.5 | 0.1 | 15.7 |
| 2004 | 0.0 | 0.0 | 15.1 | 27.9 | 3.5 | 4.1 | 1.0 | 0.5 | 0.1 | 0.0 | 52.3 |
| 2005 | 0.0 | 0.0 | 6.6 | 17.4 | 12.7 | 2.6 | 3.8 | 1.1 | 0.4 | 0.3 | 44.8 |
| 2006 | 0.0 | 0.1 | 3.5 | 8.8 | 14.0 | 22.4 | 5.1 | 5.3 | 2.1 | 1.0 | 62.2 |
| 2007 | 0.0 | 0.0 | 0.1 | 2.6 | 1.3 | 0.6 | 0.8 | 0.4 | 0.5 | 0.2 | 6.3 |
| 2008 | 0.0 | 0.0 | 0.1 | 0.1 | 0.2 | 0.1 | 0.1 | 0.2 | 0.0 | 0.0 | 0.7 |

Table 2.2.9: Catch at age (numbers in millions) of North Sea Autumn Spawners taken in IIIIa, and transfered to the assessment of NSAS, 1993-2008. SG Rednose's revisions and revision of 2002 splitting are included.

| Year/rings | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8 +}$ | Total |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1993 | 2795 | 2033 | 238 | 27 | 8 | 4 | 3 | 2 | 1 | 5109 |
| 1994 | 482 | 1087 | 201 | 27 | 6 | 3 | 2 | 0 | 0 | 1807 |
| 1995 | 1145 | 1181 | 147 | 10 | 3 | 1 | 1 | 0 | 0 | 2487 |
| 1996 | 516 | 961 | 154 | 13 | 3 | 1 | 1 | 0 | 0 | 1649 |
| 1997 | 68 | 305 | 125 | 20 | 1 | 1 | 0 | 0 | 0 | 521 |
| 1998 | 51 | 729 | 145 | 25 | 19 | 3 | 3 | 1 | 0 | 977 |
| 1999 | 598 | 231 | 133 | 39 | 10 | 5 | 1 | 1 | 0 | 1017 |
| 2000 | 232 | 978 | 115 | 20 | 21 | 7 | 3 | 1 | 0 | 1377 |
| 2001 | 808 | 557 | 140 | 15 | 1 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 411 | 345 | 48 | 5 | 1 | 0 | 0 | 0 | 0 | 1521 |
| 2003 | 22 | 445 | 182 | 13 | 16 | 2 | 1 | 1 | 0 | 811 |
| 2004 | 88 | 71 | 180 | 21 | 6 | 10 | 2 | 2 | 1 | 0 |
| 2005 | 96 | 307 | 159 | 16 | 5 | 2 | 2 | 0 | 0 | 0 |
| 2006 | 35 | 150 | 50 | 10 | 3 | 3 | 1 | 0 | 0 | 0 |
| 2007 | 68 | 189 | 77 | 2 | 0 | 1 | 0 | 1 | 0 | 0 |
| 2008 | 86 | 87 | 72 | 2 | 0 | 0 | 0 | 0 | 0 | 580 |

Table 2.2.10: Catch at age (numbers in millions) of the total North Sea Autumn Spawning stock 1993-2008. SG Rednose's revisions and the revision of 2002 splitting are included.

| Year/rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 10280 | 4160 | 1305 | 577 | 295 | 210 | 221 | 184 | 86 | 41 | 17358 |
| 1994 | 4437 | 1890 | 1839 | 449 | 332 | 103 | 88 | 74 | 68 | 45 | 9325 |
| 1995 | 7438 | 1665 | 1444 | 817 | 232 | 119 | 55 | 41 | 69 | 29 | 11909 |
| 1996 | 2311 | 1606 | 642 | 526 | 172 | 58 | 23 | 9 | 17 | 4 | 5368 |
| 1997 | 431 | 480 | 688 | 447 | 285 | 109 | 31 | 12 | 19 | 6 | 2507 |
| 1998 | 260 | 978 | 1220 | 538 | 276 | 176 | 89 | 15 | 17 | 4 | 3572 |
| 1999 | 1566 | 304 | 616 | 1059 | 294 | 136 | 69 | 28 | 10 | 2 | 4084 |
| 2000 | 1105 | 1172 | 623 | 463 | 647 | 213 | 82 | 36 | 15 | 2 | 4358 |
| 2001 | 1833 | 614 | 806 | 477 | 274 | 312 | 89 | 37 | 17 | 2 | 4463 |
| 2002 | 730 | 835 | 553 | 903 | 284 | 133 | 161 | 46 | 33 | 7 | 3687 |
| 2003 | 369 | 617 | 1204 | 517 | 820 | 243 | 106 | 120 | 37 | 8 | 4042 |
| 2004 | 716 | 207 | 439 | 1326 | 520 | 726 | 171 | 101 | 71 | 22 | 4298 |
| 2005 | 1016 | 716 | 355 | 486 | 1318 | 480 | 576 | 115 | 108 | 39 | 5209 |
| 2006 | 879 | 222 | 401 | 311 | 465 | 999 | 253 | 249 | 63 | 44 | 3885 |
| 2007 | 621 | 236 | 219 | 412 | 283 | 308 | 628 | 147 | 132 | 23 | 3009 |
| 2008 | 798 | 235 | 332 | 185 | 199 | 137 | 118 | 215 | 74 | 43 | 2336 |

Table 2.2.11: Comparison of mean weights (kg) at age (rings) in the catch of adult herring in the North Sea (by Div.) and North Sea autumn spawners caught in Div. IIIa in 1998-2008. SG Rednose's revisions are included.

| Age (Rings) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Div. | Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |  |
| IIIa | 1998 | 0.078 | 0.118 | 0.163 | 0.180 | 0.197 | 0.179 | 0.226 | - |  |
|  | 1999 | 0.084 | 0.113 | 0.141 | 0.161 | 0.181 | 0.206 | 0.199 | - |  |
|  | 2000 | 0.076 | 0.103 | 0.162 | 0.190 | 0.184 | 0.186 | 0.177 | - |  |
|  | 2001 | 0.073 | 0.105 | 0.128 | 0.133 | 0.224 | 0.170 | 0.192 | - |  |
|  | 2002 | 0.104 | 0.126 | 0.144 | 0.164 | 0.180 | 0.180 | 0.218 | - |  |
|  | 2003 | 0.067 | 0.123 | 0.150 | 0.163 | 0.191 | 0.214 | 0.187 | - |  |
|  | 2004 | 0.070 | 0.121 | 0.141 | 0.152 | 0.170 | 0.187 | 0.178 |  |  |
|  | 2005 | 0.071 | 0.106 | 0.155 | 0.173 | 0.185 | 0.200 | 0.209 | - |  |
|  | 2006 | 0.079 | 0.117 | 0.140 | 0.186 | 0.191 | 0.216 | 0.207 | - |  |
|  | 2007 | 0.071 | 0.108 | 0.125 | 0.152 | 0.184 | 0.175 | 0.154 | - |  |
|  | 2008 | 0.087 | 0.109 | 0.139 | 0.168 | 0.176 | 0.204 | 0.198 | - |  |
| IVa(E) | 1998 | 0.114 | 0.148 | 0.171 | 0.199 | 0.219 | 0.237 | 0.269 | 0.233 |  |
|  | 1999 | 0.125 | 0.143 | 0.162 | 0.191 | 0.207 | 0.226 | 0.232 | 0.272 |  |
|  | 2000 | 0.130 | 0.154 | 0.172 | 0.195 | 0.202 | 0.218 | 0.261 | 0.256 |  |
|  | 2001 | 0.121 | 0.148 | 0.165 | 0.177 | 0.197 | 0.220 | 0.262 | 0.238 |  |
|  | 2002 | 0.130 | 0.154 | 0.167 | 0.189 | 0.198 | 0.212 | 0.229 | 0.238 |  |
|  | 2003 | 0.122 | 0.154 | 0.162 | 0.177 | 0.189 | 0.203 | 0.213 | 0.218 |  |
|  | 2004 | 0.119 | 0.133 | 0.171 | 0.185 | 0.212 | 0.192 | 0.218 | 0.252 |  |
|  | 2005 | 0.117 | 0.146 | 0.153 | 0.202 | 0.209 | 0.233 | 0.262 | 0.265 |  |
|  | 2006 | 0.125 | 0.149 | 0.164 | 0.175 | 0.214 | 0.224 | 0.229 | 0.254 |  |
|  | 2007 | 0.156 | 0.148 | 0.156 | 0.186 | 0.184 | 0.204 | 0.226 | 0.239 |  |
|  | 2008 | 0.138 | 0.173 | 0.172 | 0.174 | 0.216 | 0.210 | 0.253 | 0.266 |  |
| IVa(W) | 1998 | 0.130 | 0.170 | 0.205 | 0.244 | 0.263 | 0.270 | 0.308 | 0.314 |  |
|  | 1999 | 0.129 | 0.162 | 0.192 | 0.227 | 0.250 | 0.261 | 0.272 | 0.309 |  |
|  | 2000 | 0.127 | 0.159 | 0.187 | 0.214 | 0.237 | 0.271 | 0.293 | 0.265 |  |
|  | 2001 | 0.138 | 0.168 | 0.193 | 0.222 | 0.235 | 0.266 | 0.285 | 0.296 |  |
|  | 2002 | 0.144 | 0.161 | 0.191 | 0.211 | 0.230 | 0.242 | 0.261 | 0.263 |  |
|  | 2003 | 0.130 | 0.167 | 0.184 | 0.202 | 0.224 | 0.237 | 0.259 | 0.276 |  |
|  | 2004 | 0.131 | 0.155 | 0.193 | 0.220 | 0.242 | 0.251 | 0.246 | 0.299 |  |
|  | 2005 | 0.122 | 0.158 | 0.174 | 0.213 | 0.229 | 0.245 | 0.275 | 0.267 |  |
|  | 2006 | 0.145 | 0.156 | 0.180 | 0.193 | 0.230 | 0.251 | 0.247 | 0.286 |  |
|  | 2007 | 0.150 | 0.156 | 0.166 | 0.196 | 0.191 | 0.227 | 0.241 | 0.264 |  |
|  | 2008 | 0.142 | 0.187 | 0.187 | 0.188 | 0.230 | 0.219 | 0.262 | 0.281 |  |
| IVb | 1998 | 0.117 | 0.162 | 0.203 | 0.216 | 0.243 | 0.218 | 0.311 | 0.307 |  |
|  | 1999 | 0.118 | 0.148 | 0.154 | 0.207 | 0.226 | 0.209 | 0.287 | 0.345 |  |
|  | 2000 | 0.118 | 0.173 | 0.194 | 0.224 | 0.229 | 0.251 | 0.240 | 0.268 |  |
|  | 2001 | 0.105 | 0.150 | 0.176 | 0.188 | 0.199 | 0.206 | 0.244 | 0.275 |  |
|  | 2002 | 0.086 | 0.149 | 0.161 | 0.206 | 0.214 | 0.189 | 0.270 | 0.241 |  |
|  | 2003 | 0.098 | 0.161 | 0.178 | 0.195 | 0.214 | 0.214 | 0.222 | 0.281 |  |
|  | 2004 | 0.118 | 0.143 | 0.186 | 0.214 | 0.234 | 0.239 | 0.297 | 0.308 |  |
|  | 2005 | 0.132 | 0.172 | 0.187 | 0.217 | 0.220 | 0.245 | 0.253 | 0.252 |  |
|  | 2006 | 0.097 | 0.141 | 0.172 | 0.183 | 0.202 | 0.220 | 0.232 | 0.239 |  |
|  | 2007 | 0.145 | 0.160 | 0.180 | 0.201 | 0.210 | 0.246 | 0.234 | 0.252 |  |
|  | 2008 | 0.142 | 0.172 | 0.185 | 0.191 | 0.222 | 0.228 | 0.265 | 0.223 |  |
| IVa \& IVb | 1998 | 0.123 | 0.162 | 0.194 | 0.224 | 0.243 | 0.253 | 0.293 | 0.283 |  |
|  | 1999 | 0.124 | 0.155 | 0.179 | 0.213 | 0.236 | 0.250 | 0.264 | 0.301 |  |
|  | 2000 | 0.125 | 0.162 | 0.185 | 0.210 | 0.227 | 0.258 | 0.275 | 0.263 |  |
|  | 2001 | 0.129 | 0.156 | 0.180 | 0.202 | 0.217 | 0.242 | 0.275 | 0.285 |  |
|  | 2002 | 0.119 | 0.157 | 0.177 | 0.203 | 0.219 | 0.228 | 0.253 | 0.253 |  |
|  | 2003 | 0.113 | 0.163 | 0.178 | 0.190 | 0.210 | 0.225 | 0.239 | 0.255 |  |
|  | 2004 | 0.122 | 0.147 | 0.187 | 0.210 | 0.227 | 0.233 | 0.247 | 0.266 |  |
|  | 2005 | 0.121 | 0.157 | 0.172 | 0.212 | 0.225 | 0.242 | 0.269 | 0.265 |  |
|  | 2006 | 0.123 | 0.150 | 0.174 | 0.187 | 0.222 | 0.239 | 0.238 | 0.269 |  |
|  | 2007 | 0.149 | 0.155 | 0.165 | 0.196 | 0.192 | 0.227 | 0.238 | 0.257 |  |
|  | 2008 | 0.142 | 0.182 | 0.185 | 0.188 | 0.226 | 0.220 | 0.262 | 0.275 |  |
| $\overline{\text { IVc \& VIId }}$ | 1998 | 0.096 | 0.114 | 0.146 | 0.149 | 0.184 | 0.000 | 0.176 | - |  |
|  | 1999 | 0.116 | 0.139 | 0.159 | 0.189 | 0.198 | 0.217 | - | - |  |
|  | 2000 | 0.106 | 0.133 | 0.150 | 0.180 | 0.194 | 0.203 | - | - |  |
|  | 2001 | 0.113 | 0.138 | 0.171 | 0.167 | 0.171 | 0.168 | 0.180 | - |  |
|  | 2002 | 0.108 | 0.123 | 0.153 | 0.170 | 0.187 | 0.219 | 0.208 | - |  |
|  | 2003 | 0.103 | 0.127 | 0.144 | 0.168 | 0.176 | 0.188 | 0.200 | 0.227 |  |
|  | 2004 | 0.099 | 0.113 | 0.135 | 0.162 | 0.184 | 0.191 | 0.186 | 0.224 |  |
|  | 2005 | 0.122 | 0.132 | 0.139 | 0.170 | 0.207 | 0.228 | 0.237 | 0.245 |  |
|  | 2006 | 0.119 | 0.125 | 0.153 | 0.152 | 0.178 | 0.205 | 0.209 | 0.219 |  |
|  | 2007 | 0.129 | 0.131 | 0.154 | 0.158 | 0.173 | 0.196 | 0.209 | 0.218 |  |
|  | 2008 | 0.120 | 0.157 | 0.156 | 0.173 | 0.188 | 0.192 | 0.215 | 0.247 |  |
| Total | 1998 | 0.119 | 0.146 | 0.185 | 0.219 | 0.239 | 0.253 | 0.288 | 0.283 |  |
| North Sea | 1999 | 0.123 | 0.152 | 0.172 | 0.208 | 0.233 | 0.246 | 0.264 | 0.301 |  |
| Catch | 2000 | 0.122 | 0.159 | 0.180 | 0.202 | 0.217 | 0.247 | 0.275 | 0.263 |  |
|  | 2001 | 0.118 | 0.149 | 0.177 | 0.198 | 0.213 | 0.238 | 0.267 | 0.288 |  |
|  | 2002 | 0.118 | 0.153 | 0.170 | 0.199 | 0.214 | 0.228 | 0.250 | 0.252 |  |
|  | 2003 | 0.104 | 0.158 | 0.174 | 0.184 | 0.205 | 0.222 | 0.232 | 0.256 |  |
|  | 2004 | 0.100 | 0.138 | 0.183 | 0.201 | 0.216 | 0.228 | 0.246 | 0.272 |  |
|  | 2005 | 0.099 | 0.153 | 0.166 | 0.208 | 0.223 | 0.240 | 0.257 | 0.278 |  |
|  | 2006 | 0.122 | 0.145 | 0.172 | 0.181 | 0.220 | 0.237 | 0.235 | 0.262 |  |
|  | 2007 | 0.149 | 0.152 | 0.164 | 0.194 | 0.190 | 0.224 | 0.235 | 0.252 |  |
|  | 2008 | 0.141 | 0.180 | 0.181 | 0.183 | 0.216 | 0.216 | 0.256 | 0.273 |  |

in the danish catches and new information of misreportings from the UK.

Table 2.2.12: Sampling of commercial landings of herring in the North Sea (Div. IV and VIId) in 2008 by quarter. Sampled catch means the proportion of the reported catch to which sampling was applied. It is limited by $100 \%$ but might exceed the official landings due to sampling of discards, unallocated and misreported catches. It is not possible to judge the quality of the sampling by this figure alone. Note that only one nation sampled their by-catches in the industrial fishery (Denmark, fleet B). Metiers are each reported combination of nation/fleet/area/quarter.

| Country (fleet) | Quarter | No of metiers | Metiers sampled | Sampled Catch \% | Official Catch | No. of samples | $\begin{array}{r} \text { No. fish } \\ \text { aged } \end{array}$ | No. fish measured | >1 sample per 1 kt catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark (A) | 1 | 3 | 0 | 0\% | 8776 | 0 | 0 | 0 | n |
|  | 2 | 3 | 1 | 92\% | 2109 | 2 | 55 | 292 | n |
|  | 3 | 3 | 3 | 100\% | 34864 | 19 | 504 | 2730 | n |
|  | 4 | 4 | 1 | 52\% | 8509 | 1 | 25 | 141 | n |
| total |  | 13 | 5 | 76\% | 54258 | 22 | 584 | 3163 | n |
| Denmark (B) | 1 | 3 | 1 | 61\% | 17 | 3 | 16 | 16 | y |
|  | 2 | 1 | 0 | 0\% | 195 | 9 | 0 | 11 | y |
|  | 3 | 2 | 1 | 93\% | 4262 | 9 | 395 | 405 | y |
|  | 4 | 3 | 1 | 76\% | 4133 | 4 | 71 | 72 | y |
| total |  | 9 | 3 | 83\% | 8606 | 25 | 482 | 504 | y |
| England and Wales* | 1 | 2 | 0 | 0\% | 179 | 0 | 0 | 0 | n |
|  | 2 | 2 | 1 | 100\% | 872 | 5 | 125 | 629 | y |
|  | 3 | 3 | 2 | 100\% | 7893 | 11 | 275 | 1274 | y |
|  | 4 | 4 | 0 | 0\% | 2772 | 0 | 0 | 0 | n |
| total |  | 11 | 3 | 75\% | 11717 | 16 | 400 | 1903 | y |
| Faroe | 1 | 2 | 0 | 0\% | 600 | 0 | 0 | 0 | n |
| Island | 4 | 2 | 0 | 0\% | 1414 | 0 | 0 | 0 | n |
| total |  | 4 | 0 | 0\% | 2014 | 0 | 0 | 0 | n |
| France | 1 | 3 | 0 | 0\% | 619 | 0 | 0 | 0 | n |
|  | 2 | 4 | 0 | 0\% | 1156 | 0 | 0 | 0 | n |
|  | 3 | 4 | 0 | 0\% | 20925 | 0 | 0 | 0 | n |
|  | 4 | 3 | 0 | 0\% | 7647 | 0 | 0 | 0 | n |
| total |  | 14 | 0 | 0\% | 30346 | 0 | 0 | 0 | n |
| Germany | 3 | 2 | 0 | 0\% | 2145 | 0 | 0 | 0 | n |
|  | 4 | 4 | 4 | 100\% | 5965 | 27 | 2284 | 11878 | y |
| total |  | 6 | 4 | 74\% | 8109 | 27 | 2284 | 11878 | y |
| Netherlands | 1 | 2 | 2 | 100\% | 1391 | 17 | 425 | 3014 | y |
|  | 2 | 1 | 1 | 100\% | 4470 | 17 | 425 | 2626 | y |
|  | 3 | 3 | 1 | 100\% | 8325 | 29 | 725 | 2891 | y |
|  | 4 | 4 | 1 | 100\% | 8936 | 4 | 100 | 509 | n |
| total |  | 10 | 5 | 100\% | 23122 | 67 | 1675 | 9040 | y |
| Northern Irelan total | 1 | 1 | 0 | 0\% | 331 | 0 | 0 | 0 |  |
|  |  | 1 | 0 | 0\% | 331 | 0 | 0 | 0 | n |
| Norway | 1 | 1 | 0 | 0\% | 489 | 0 | 0 | 0 | n |
|  | 2 | 3 | 2 | 99\% | 34231 | 17 | 584 | 1700 | n |
|  | 3 | 3 | 0 | 0\% | 3695 | 0 | 0 | 0 | n |
|  | 4 | 3 | 3 | 100\% | 20906 | 8 | 313 | 572 | n |
| total |  | 10 | 5 | 93\% | 59321 | 25 | 897 | 2272 | n |
| Scotland | 1 | 1 | 0 | 0\% | 895 | 0 | 0 | 0 | n |
|  | 2 | 2 | 1 | 86\% | 1243 | 7 | 355 | 1270 | y |
|  | 3 | 2 | 2 | 100\% | 12735 | 26 | 1875 | 5823 | y |
|  | 4 | 3 | 1 | 94\% | 1183 | 2 | 111 | 379 | $y$ |
| total |  | 8 | 4 | 100\% | 16056 | 35 | 2341 | 7472 | y |
| Sweden | 2 | 2 | 0 | 0\% | 6625 | 0 | 0 | 0 | n |
|  | 3 | 3 | 0 | 0\% | 5085 | 0 | 0 | 0 | n |
|  | 4 | 2 | 0 | 0\% | 2305 | 0 | 0 | 0 | n |
| total |  | 5 | 0 | 0\% | 7390 | 0 | 0 | 0 | n |
|  |  | 93 | 29 | 76\% | 227895 | 217 | 8663 | 36232 | y |
| grand total | 1 | 18 | 3 | 24\% | 13297 | 20 | 441 | 3030 | n |
| Period total 2 |  | 18 | 6 | 93\% | 50900 | 57 | 1544 | 6528 | y |
| Period total 3 |  | 25 | 9 | 77\% | 99928 | 94 | 3774 | 13123 | y |
| Period total 4 |  | 32 | 11 | 72\% | 63770 | 46 | 2904 | 13551 | n |
| Total for stock 2008 |  | 93 | 29 | 76\% | 227895 | 217 | 8663 | 36232 | n |
| Human Cons. only |  | 84 | 26 | 76\% | 219290 | 192 | 8181 | 35728 | n |
| Total for stock 2006 |  | 107 | 39 | 79\% | 490362 | 404 | 23581 | 65536 | n |
| Total for stock 2007 |  | 100 | 30 | 86\% | 361114 | 335 | 10342 | 54639 | n |
| Human Cons. only 2007 |  | 91 | 27 | 85\% | 354017 | 318 | 10194 | 54310 | n |

[^0]Herring catches 2008, 1st Quarter


Figure 2.1.1a: : Herring catches in the 1st quarter in the North Sea, in Div VIId, Div IIIa, SD 22 and SD 24 (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available).

Herring catches 2008, 2nd Quarter


Figure 2.1.1b: Herring catches in the 2nd quarter in the North Sea, in Div VIId, Div IIIa, SD 22 and SD 24 (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available).

## Herring catches 2008, 3rd Quarter



Figure 2.1.1c: Herring catches in the 3rd quarter in the North Sea, in Div VIId, Div IIIa, SD 22 and SD 24 (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available).

## Herring catches 2008, 4th Quarter



Figure 2.1.1d: Herring catches in the 4th quarter in the North Sea, in Div VIId, Div IIIa, SD 22 and SD 24 (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available).

Herring catches 2008, All Quarters


Figure 2.1.1e: Herring catches in all quarters in the North Sea, in Div VIId, Div IIIa, SD 22 and SD 24 (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available).



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Figure 2.2.1: Proportions of age groups (numbers) in the total catch of herring in the North Sea (upper, 1960-2008, and middle panel, 1980-2008), and in the total catch of North Sea autumn spawners in 2008 (lower panel).


Figure 2.2.2: Mean vertebrae counts of 2 (upper number), 3 (middle) and $4+$ herring (lower) in the North Sea and Div. IIIa as obtained by Norwegian sampling in the $2^{\text {nd }}$ and $3^{\text {rd }}$ quarter 2008. The transfer area (Western Baltic spring spawners transferred to the assessment of IIIa herring) is indicated.

### 2.3 Fishery Independent Information

### 2.3.1 Acoustic Surveys in the North Sea, West of Scotland VIa(N) and the Malin Shelf area in July 2008

Eight surveys were carried out during late June and July 2008 covering most of the continental shelf in the North Sea, West of Scotland and the Malin Shelf. The individual surveys and the survey methods are given in the report of the Planning Group of International Pelagic Surveys (PGIPS; ICES, 2009/LRC:02). The vessels, areas and dates of cruises are given in Table 2.3.1.1 and in Figure 2.3.1.1.

The data has been combined to provide an overall estimate of numbers-at-age, maturity ogive and mean weights-at-age. These have been calculated as weighted means of individual survey estimates by ICES statistical rectangle. The weighting applied is proportional to the survey track for each vessel that has been covered in each statistical rectangle.

The estimate of North Sea autumn spawning herring spawning stock is higher than the previous year, at 1.8 million tonnes and 9.5 million herring (Table 2.3.1.2). The survey indicates that the strong 2000 year class of herring still persists in the population. Growth of the 2000 year class seems still to be slower than average, with individuals of this year class having a lower mean size and mean weight of those fish which are one year younger (the 2001 year class).

The spatial distribution of mature and immature autumn spawning herring is shown in Figures 2.3.1.2 and 2.3.1.3 respectively. Adult herring in the North Sea are concentrated in northern areas close to the Fladen grounds.

The time series of abundance for North Sea autumn spawners are given in Table 2.3.1.3.

Table 2.3.1.1: Pelagic Acoustic Surveys. Vessels, areas and cruise dates in 2008.

| Vessel | Period | Area | Rectangles |
| :---: | :---: | :---: | :---: |
| Corystes (NIR) | $\begin{aligned} & 05 \text { July - } 11 \\ & \text { July } \end{aligned}$ | Clyde/ North Channel | 40E3-E5, 39E4-E5,38E4 |
| Celtic <br> Explorer <br> (IR) | $\begin{aligned} & 28 \text { June - } 14 \\ & \text { July } \end{aligned}$ | $52^{\circ} 30^{\prime}-56^{\circ} \mathrm{N}, 12^{\circ}-6^{\circ} \mathrm{W}$ | 34D9-E0, 35D9-E0, 36D9-E0, 37D9-E1, 38D9E1, 39E0-E3, 40E1-E3 |
| Charter <br> west Sco (SCO) | $\begin{aligned} & 30 \text { June - } 19 \\ & \text { July } \end{aligned}$ | $\begin{aligned} & 55^{\circ} 30^{\prime}-60^{\circ} 30^{\prime} \mathrm{N}, 4^{\circ}- \\ & 10^{\circ} \mathrm{W} \end{aligned}$ | 41E0-E3, 42E0-E3, 43E0-E3, 44E0-E3, 45E0E4, 46E2-E5, 47E2-E5, 48E4-E5, 49E5 |
| Johan Hjort (NOR) | $\begin{aligned} & 01 \text { July - } 31 \\ & \text { July } \end{aligned}$ | $56^{\circ} 30^{\prime}-62^{\circ} \mathrm{N}, 2^{\circ}-6^{\circ} \mathrm{E}$ | $\begin{aligned} & \text { 42F2-F5, 43F2-F5, 44F2-F5, 45F2-F5, 46F2-F4, } \\ & \text { 47F2-F4, 48F2-F4, 49F2-F4, 50F2-F4, 51F2-F4, } \\ & \text { 52F2-F4 } \end{aligned}$ |
| Scotia <br> (SCO) | $\begin{aligned} & 28 \text { June - } 18 \\ & \text { July } \end{aligned}$ | $58^{\circ} 30^{\prime}-62^{\circ} \mathrm{N}, 4^{\circ} \mathrm{W}-2^{\circ} \mathrm{E}$ | $\begin{aligned} & 46 \mathrm{E} 6-\mathrm{F} 1,47 \mathrm{E} 6-\mathrm{F} 1,48 \mathrm{E} 6-\mathrm{F} 1,49 \mathrm{E} 6-\mathrm{F} 1,50 \mathrm{E} 7-\mathrm{F} 1, \\ & 51 \mathrm{E} 8-\mathrm{F} 1,52 \mathrm{E}-\mathrm{F} 1 \end{aligned}$ |
| Tridens (NED) | $\begin{aligned} & 23 \text { June - } 18 \\ & \text { July } \end{aligned}$ | $\begin{aligned} & 54^{\circ}-58^{\circ} 30^{\prime} \mathrm{N}, 4^{\circ} \mathrm{W}- \\ & 2^{\circ} / 6^{\circ} \mathrm{E} \end{aligned}$ | 37E9-F1, 38E8-F1, 39E8-F1, 40E8-F5, 41E7-F5, 42E7-F1, 43E7-F1, 44E6-F1, 45E6-F1 |
| Solea (GER) DBFH | $\begin{aligned} & 26 \text { June - } 16 \\ & \text { July } \end{aligned}$ | $52^{\circ}-56^{\circ} \mathrm{N}$, Eng to Den/Ger coasts | 33F1-F4, 34F2-F4, 35F2-F4, 36F0-F7, 37F2-F8, 38F2-F7, 39F2-F7, 40F6-F7 |
| Dana (DEN) OXBH | $\begin{aligned} & 26 \text { June - } 07 \\ & \text { July } \end{aligned}$ | Kattegat and North of $56^{\circ} \mathrm{N}$, east of $6^{\circ} \mathrm{E}$ | 41 F6-F7, 41G1-G2, 42F6-F7, 42G0-G2, 43F6G1, 44F6-G1, 45F8-G1, 46F9-G0 |

Table 2.3.1.2: Total numbers (millions of fish) and biomass (thousands of tonnes) of North Sea autumn spawning herring in the area surveyed in the pelagic acoustic surveys July 2008, with mean weights and mean lengths by age ring.

| Age ( ring) | Numbers | Biomass | Maturity | weight(g) | Length (cm) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 6,870 | 60 | 0.00 | 8.7 | 10.5 |
| 1 | 3,714 | 232 | 0.05 | 62.4 | 19.2 |
| 2 | 2,853 | 403 | 0.86 | 141.4 | 25.0 |
| 3 | 1,709 | 307 | 0.98 | 179.7 | 26.8 |
| 4 | 1,485 | 272 | 0.99 | 183.3 | 27.0 |
| 5 | 809 | 157 | 1.00 | 194.4 | 27.5 |
| 6 | 712 | 164 | 1.00 | 229.9 | 28.7 |
| 7 | 1,749 | 380 | 1.00 | 217.4 | 28.4 |
| 8 | 185 | 50 | 1.00 | 267.9 | 29.7 |
| 9+ | 270 | 76 | 1.00 | 282.3 | 30.2 |
| IMMATURE | 10,841 | 317 |  | 29.2 | 13.8 |
| Mature | 9,514 | 1,784 |  | 187.5 | 27.0 |
| Total | 20,355 | 2,100 | 0.47 | 103.2 | 20.0 |

Table 2.3.1.3: Estimates of North Sea autumn spawners (millions) at age from acoustic surveys, 1985-2008. For 1985-1986 the estimates are the sum of those from the Division IVa summer survey, the Division IVb autumn survey, and the Divisions IVc, VIId winter survey. The 1987 to 2008 estimates are from the summer survey in Divisions IVa,b and IIIa excluding estimates of Division IIIa/Baltic spring spawners. For 1999 and 2000 the Kattegat was excluded from the results because it was not surveyed.

| Years / <br> Age (rings) | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | Total | $\begin{gathered} \text { SSB } \\ (‘ \text { '000T) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 726 | 2,789 | 1,433 | 323 | 113 | 41 | 17 | 23 | 19 | 5,484 | 697 |
| 1986 | 1,639 | 3,206 | 1,637 | 833 | 135 | 36 | 24 | 6 | 8 | 7,542 | 942 |
| 1987 | 13,736 | 4,303 | 955 | 657 | 368 | 77 | 38 | 11 | 20 | 20,165 | 817 |
| 1988 | 6,431 | 4,202 | 1,732 | 528 | 349 | 174 | 43 | 23 | 14 | 13,496 | 897 |
| 1989 | 6,333 | 3,726 | 3,751 | 1,612 | 488 | 281 | 120 | 44 | 22 | 16,377 | 1,637 |
| 1990 | 6,249 | 2,971 | 3,530 | 3,370 | 1,349 | 395 | 211 | 134 | 43 | 18,262 | 2,174 |
| 1991 | 3,182 | 2,834 | 1,501 | 2,102 | 1,984 | 748 | 262 | 112 | 56 | 12,781 | 1,874 |
| 1992 | 6,351 | 4,179 | 1,633 | 1,397 | 1,510 | 1,311 | 474 | 155 | 163 | 17,173 | 1,545 |
| 1993 | 10,399 | 3,710 | 1,855 | 909 | 795 | 788 | 546 | 178 | 116 | 19,326 | 1,216 |
| 1994 | 3,646 | 3,280 | 957 | 429 | 363 | 321 | 238 | 220 | 132 | 13,003 | 1,035 |
| 1995 | 4,202 | 3,799 | 2,056 | 656 | 272 | 175 | 135 | 110 | 84 | 11,220 | 1,082 |
| 1996 | 6,198 | 4,557 | 2,824 | 1,087 | 311 | 99 | 83 | 133 | 206 | 18,786 | 1,446 |
| 1997 | 9,416 | 6,363 | 3,287 | 1,696 | 692 | 259 | 79 | 78 | 158 | 22,028 | 1,780 |
| 1998 | 4,449 | 5,747 | 2,520 | 1,625 | 982 | 445 | 170 | 45 | 121 | 16,104 | 1,792 |
| 1999 | 5,087 | 3,078 | 4,725 | 1,116 | 506 | 314 | 139 | 54 | 87 | 15,107 | 1,534 |
| 2000 | 24,735 | 2,922 | 2,156 | 3,139 | 1,006 | 483 | 266 | 120 | 97 | 34,928 | 1,833 |
| 2001 | 6,837 | 12,290 | 3,083 | 1,462 | 1,676 | 450 | 170 | 98 | 59 | 26,124 | 2,622 |
| 2002 | 23,055 | 4,875 | 8,220 | 1,390 | 795 | 1,031 | 244 | 121 | 150 | 39,881 | 2,948 |
| 2003 | 9,829 | 18,949 | 3,081 | 4,189 | 675 | 495 | 568 | 146 | 178 | 38,110 | 2,999 |
| 2004 | 5,183 | 3,415 | 9,191 | 2,167 | 2,590 | 317 | 328 | 342 | 186 | 23,722 | 2,584 |
| 2005 | 3,113 | 1,890 | 3,436 | 5,609 | 1,211 | 1,172 | 140 | 127 | 107 | 16,805 | 1,868 |
| 2006 | 6,823 | 3,772 | 1,997 | 2,098 | 4,175 | 618 | 562 | 84 | 70 | 20,199 | 2,130 |
| 2007 | 6,261 | 2,750 | 1,848 | 898 | 806 | 1,323 | 243 | 152 | 65 | 14,346 | 1,203 |
| 2008 | 3,714 | 2,853 | 1,709 | 1,485 | 809 | 712 | 1,749 | 185 | 270 | 20,355 | 1,784 |



Figure 2.3.1.1: Survey area coverage in the pelagic acoustic surveys in 2008, by rectangle and nation (IR = Celtic Explorer; NIR = Corystes; WSC = West of Scotland charter vessel; SCO = Scotia; NOR = Johan Hjort; DK = Dana; NL = Tridens; GER = Solea). Multi-coloured rectangles indicate overlapping coverage by two or more nations (e.g. 40E1-40E3).


Figure 2.3.1.2: Biomass of mature autumn spawning herring from the combined acoustic survey in June - July 2008 (maximum grid density = 200000 t).


Figure 2.3.1.3: Biomass of immature autumn spawning herring from the combined acoustic survey in June - July 2008 (maximum grid density $=56000$ t).

### 2.3.2 Larvae Surveys in the North Sea 2008

Seven larvae surveys were conducted between September 2008 and January 2009 (Table 2.3.2.1). The survey effort in numbers of samples taken and vessel days in the surveys was comparable to previous years (Table 2.3.2.2).

Large numbers of newly hatched herring larvae were obtained from all areas observed, with the only exception of the Buchan area. When compared to previous years, herring larvae abundance has decreased in the Buchan area, while the Larvae Abundance Indices (LAI) estimated for the Orkney/Shetland area, the central North Sea area and the southern North Sea have increased (Tab. 2.3.2.3, Figure 2.3.2.1). The LAI in the Central North Sea revealed the highest level observed since 2003. However, this area is well known for large annual variabilities both in larvae abundance and survey effort.

The MLAI for the whole North Sea derived from the larvae surveys in period 2008/2009 indicate that the SSB has increased when compared to last year's WG estimate (Figure 2.3.2.2). The updated MLAI time-series is shown in Table 2.3.2.3.

Detailed information on survey coverage and effort in the North Sea are given in the Report of the herring larvae surveys in the North Sea (Rohlf \& Gröger, WD 12).

Table 2.3.2.1: North Sea autumn spawners, Herring Larvae Surveys. Fortnightly time periods sampled and survey effort in 2008/2009.

NL - Netherlands, FRG - Federal Republic of Germany

| Area | Time period | Samples available | Vessel days | Nation | Coverage |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Orkney/Shetland | 01-15 Sep. | 78 | 7 | GER | Total |
|  | 16-30 Sep. | 78 | 5 | GER | Total |
| Buchan | $01-15$ Sep. | None |  |  |  |
|  | 16-30 Sep. | 80 | 5 | NL | Total |
| Central North | $01-15$ Sep. | None |  |  |  |
| Sea | 16-30 Sep. | 63 | 4 | NL | Partly |
|  | 01-15 Oct. | None |  |  |  |
| Southern North | 16-31 Dec. | 78 | 4 | NL | Total |
| Sea | 01-15 Jan. | 21 | 2 | GER | Partly |
|  | 16-31 Jan. | 83 | 4 | NL | Total |

Table 2.3.2.2: North Sea autumn spawners, Herring Larvae Surveys. Number of samples taken and sampling effort in Orkney/Shetland, Buchan, Central North Sea and Southern North Sea by year

| Year | SAMPLES | Vessel-days (SAMPLING) |
| :--- | :---: | :---: |
| $1988 / 89$ | 1355 | 98 |
| $1989 / 90$ | 1300 | 96 |
| $1990 / 91$ | 634 | 49 |
| $1991 / 92$ | 738 | 51 |
| $1992 / 93$ | 498 | 31 |
| $1993 / 94$ | 491 | 34 |
| $1994 / 95$ | 450 | 33 |
| $1995 / 96$ | 421 | 26 |
| $1996 / 97$ | 469 | 32 |
| $1997 / 98$ | 456 | 29 |
| $1998 / 99$ | 531 | 37 |
| $1999 / 00$ | 645 | 38 |
| $2000 / 01$ | 696 | 53 |
| $2001 / 02$ | 534 | 32 |
| $2002 / 03$ | 533 | 35 |
| $2003 / 04$ | 568 | 35 |
| $2004 / 05$ | 483 | 33 |
| $2005 / 06$ | 543 | 36 |
| $2006 / 07$ | 568 | 35 |
| $2007 / 08$ | 495 | 34 |
| $2008 / 09$ | 481 | 31 |

Table 2.3.2.3: North Sea autumn spawners. Estimated abundances of herring larvae $<\mathbf{1 0} \mathrm{mm}$ long ( $<11 \mathrm{~mm}$ for the SNS), by standard sampling area and time periods. The number of larvae are expressed as mean number per ICES rectangle * $10^{9}$

|  | Orkney/ <br> SHETLAND |  | Buchan |  | Central North Sea |  |  | Southern North Sea |  |  | MLAI AsSess |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Period | $\begin{gathered} 1-15 \\ \text { SEP. } \end{gathered}$ | 16-30 <br> Sep. | $\begin{gathered} 1-15 \\ \text { SEP. } \end{gathered}$ | $\begin{gathered} 16- \\ 30 \\ \text { SEP. } \end{gathered}$ | $\begin{gathered} 1-15 \\ \text { SEP. } \end{gathered}$ | $\begin{gathered} 16- \\ 30 \\ \text { SEP. } \end{gathered}$ | $\begin{aligned} & 1-15 \\ & \text { Ост. } \end{aligned}$ | $\begin{gathered} 16- \\ 31 \\ \text { DEC. } \end{gathered}$ | $1-15$ <br> JAN. | $\begin{gathered} 16- \\ 31 \\ \text { JAN. } \end{gathered}$ |  |
| 1972 | 1133 | 4583 | 30 |  | 165 | 88 | 134 | 2 | 46 |  |  |
| 1973 | 2029 | 822 | 3 | 4 | 492 | 830 | 1213 |  |  | 1 | 13.182 |
| 1974 | 758 | 421 | 101 | 284 | 81 |  | 1184 |  | 10 |  | 7.943 |
| 1975 | 371 | 50 | 312 |  |  | 90 | 77 | 1 | 2 |  | 2.819 |
| 1976 | 545 | 81 |  | 1 | 64 | 108 |  |  | 3 |  | 2.494 |
| 1977 | 1133 | 221 | 124 | 32 | 520 | 262 | 89 | 1 |  |  | 6.151 |
| 1978 | 3047 | 50 |  | 162 | 1406 | 81 | 269 | 33 | 3 |  | 7.427 |
| 1979 | 2882 | 2362 | 197 | 10 | 662 | 131 | 507 |  | 111 | 89 | 14.363 |
| 1980 | 3534 | 720 | 21 | 1 | 317 | 188 | 9 | 247 | 129 | 40 | 9.771 |
| 1981 | 3667 | 277 | 3 | 12 | 903 | 235 | 119 | 1456 |  | 70 | 14.337 |
| 1982 | 2353 | 1116 | 340 | 257 | 86 | 64 | 1077 | 710 | 275 | 54 | 20.891 |
| 1983 | 2579 | 812 | 3647 | 768 | 1459 | 281 | 63 | 71 | 243 | 58 | 26.804 |
| 1984 | 1795 | 1912 | 2327 | 1853 | 688 | 2404 | 824 | 523 | 185 | 39 | 48.366 |
| 1985 | 5632 | 3432 | 2521 | 1812 | 130 | 13039 | 1794 | 1851 | 407 | 38 | 73.818 |
| 1986 | 3529 | 1842 | 3278 | 341 | 1611 | 6112 | 188 | 780 | 123 | 18 | 38.444 |
| 1987 | 7409 | 1848 | 2551 | 670 | 799 | 4927 | 1992 | 934 | 297 | 146 | 67.690 |
| 1988 | 7538 | 8832 | 6812 | 5248 | 5533 | 3808 | 1960 | 1679 | 162 | 112 | 134.382 |
| 1989 | 11477 | 5725 | 5879 | 692 | 1442 | 5010 | 2364 | 1514 | 2120 | 512 | 131.732 |
| 1990 |  | 10144 | 4590 | 2045 | 19955 | 1239 | 975 | 2552 | 1204 |  | 171.592 |
| 1991 | 1021 | 2397 |  | 2032 | 4823 | 2110 | 1249 | 4400 | 873 |  | 90.332 |
| 1992 | 189 | 4917 |  | 822 | 10 | 165 | 163 | 176 | 1616 |  | 42.147 |
| 1993 |  | 66 |  | 174 |  | 685 | 85 | 1358 | 1103 |  | 30.069 |
| 1994 | 26 | 1179 |  |  |  | 1464 | 44 | 537 | 595 |  | 20.798 |
| 1995 |  | 8688 |  |  |  |  | 43 | 74 | 230 | 164 | 22.353 |
| 1996 |  | 809 |  | 184 |  | 564 |  | 337 | 675 | 691 | 43.983 |
| 1997 |  | 3611 |  | 23 |  |  |  | 9374 | 918 | 355 | 56.462 |
| 1998 |  | 8528 |  | 1490 | 205 | 66 |  | 1522 | 953 | 170 | 72.912 |
| 1999 |  | 4064 |  | 185 |  | 134 | 181 | 804 | 1260 | 344 | 60.531 |
| 2000 |  | 3352 | 28 | 83 |  | 376 |  | 7346 | 338 | 106 | 40.441 |
| 2001 |  | 11918 |  | 164 |  | 1604 |  | 971 | 5531 | 909 | 129.562 |
| 2002 |  | 6669 |  | 1038 |  |  | 3291 | 2008 | 260 | 925 | 109.899 |
| 2003 |  | 3199 |  | 2263 |  | 12018 | 3277 | 12048 | 3109 | 1116 | 267.813 |
| 2004 |  | 7055 |  | 3884 |  | 5545 |  | 7055 | 2052 | 4175 | 321.660 |
| 2005 |  | 3380 |  | 1364 |  | 5614 |  | 498 | 3999 | 4822 | 192.265 |
| 2006 | 6311 | 2312 |  | 280 |  | 2259 |  | 10858 | 2700 | 2106 | 117.856 |
| 2007 |  | 1753 |  | 1304 |  | 291 |  | 4443 | 2439 | 3854 | 173.003 |
| 2008 | 4978 | 6875 |  | 533 |  | 11201 |  | 8426 | 2317 | 4008 | 181.746 |



Figure 2.3.2.1: North Sea autumn spawners. Larval Abundance Index time-series for a collection of areas and sampling periods ( $B=$ Orkney/Shetland $1^{\text {st }}$ and $2^{\text {nd }}$ fortnight, $C=$ Buchan $2^{\text {nd }}$ fortnight, $D=$ Central North Sea $2^{\text {nd }}$ fortnight, $E=$ Southern North Sea all 3 fortnights).


Figure 2.3.2.2: North Sea autumn spawners. Comparison of spawning stock size estimates (x axis) and year effects (y axis) when fitting small larvae abundances to SSB using the multiplicative model (with regression line and $95 \%$ confidence limits). The SSB estimate is derived from the ICA-output.

### 2.3.3 International Bottom Trawl Survey (IBTS)

The International Bottom Trawl Survey (IBTS) started out as a young herring fish survey in 1966 with the objective of obtaining annual recruitment indices (abundance of 1-ringers in $1^{\text {st }}$ quarter) for the combined North Sea herring stock. The survey has been carried out every year since, and presently it provides recruitment indices not only for herring, but for demersal species as well. Examinations of the catch of adult herring during the $1^{\text {st }}$ quarter IBTS have shown that this catch also indicates abundances of $2-5+$ herring. Further, sampling for large herring larvae (0-ringers) is carried out at night-time during the IBTS $1^{\text {st }}$ quarter using a fine-meshed 2 metre ring net (MIK). Hence, the sampling during IBTS affords an extended series of herring abundance indices ( 0 to $5+$ ringers).

### 2.3.3.1 The index of 0-ringer abundance

The total abundance of 0-ringers in the survey area is used as recruitment index for the stock. This year's 0-ringer index is based on 641 depth-integrated hauls with a 2 metre ring-net (MIK). Index values are calculated as described in the WG report of 1996 (ICES 1996/ACFM:10). The series of estimates is shown in Table 2.3.3.1, the new index value of 0-ringer abundance of the 2008 year class is estimated at 95.8.

The index indicates a significant increase in recruitment from last year's estimate, which was outstandingly low, but was one in a series of poor recruitments starting from the 2002 year class. The 0 -ringers which are included in the index were predominantly distributed off the Scottish coast, and in the northern areas (Figure 2.3.3.1). Compared to the preceding two year classes, the 0-ringers from this year class is distributed further from the Scottish coast and further to the north. A large concentration was found centrally in the North Sea, and concentrations were also seen in the Skagerrak/Kattegat. Downs herring larvae were apparent from MIK catches in the area of the English Channel, however, due to their small size (many below 12 mm mean length) most of these will not contribute to the recruitment index at a scale comparable to estimates based on larger larvae (> 20 mm ). Hence, these small larvae are not included in the standard procedure of index estimation (see ICES 1996 /ACFM:10). As for last year's index estimation the WG investigated changes in 0 -ringer estimates, when including the catches of small Downs larvae, but accounting for a daily mortality rate of 0.1 until these reached the 20 mm length. This procedure led to only slight increase in this year's index estimate, which indicates a relatively minor bias from the exclusion of this group from the present index estimation. To further investigate the influence of such changes in estimation procedures, a testseries of indices, which included the small Downs larvae, was calculated for the period 1992-09 (Table 2.3.3.2) and exploratory assessment runs were carried out. The outcome of this investigation is described in section 2.10.

The long term trend in the distributional patterns of 0-ringers is apparent from the changes in absolute and relative abundance of 0-ringers in the western part of the North Sea, as illustrated in Figure 2.3.3.2. In this figure the relative abundance is given as the number of 0-ringers in the area west of $2^{\circ} \mathrm{E}$ relative to the total number of 0 -ringers in the given year class. Since the year class 1982, when the relative abundance was $25 \%$, a general increase in abundance has been seen for the western part. In the last decade, the majority of 0-ringers has been distributed in this area, and the calculated relative abundance of $55 \%$ for the present year class is in accordance with the long term trend.

### 2.3.3.2 The indices of $1-5+$ ringer herring abundances

## 1-ringer abundance

The 1-ringer index of recruitment is based on trawl catches in the entire survey area. The time series for year classes 1977 to 2007 are shown in Table 2.3.3.3. This year's estimate of the 2007 year class strength indicates a recruitment in the order of the long term mean ( $84 \%$ ) which is a significant increase from the low recruitments estimated for the year classes 2002-2005. However, it should be noted that the index is strongly influenced by outstandingly high trawl catches in the Kattegat. The 1-ringers caught in Kattegat in 2009 are very small in size, and otolith examination of spawning origin indicates that a large fraction of these 1-ringers is from the Western Baltic Spring Spawner stock. If the fraction of potential WBSS 1-ringers is excluded in the index calculation, the index will be reduced to approx $70 \%$ of the present value. . To further investigate the influence of such changes in estimation procedures, a testseries of 1-ringer indices, which excluded the WBSS 1-ringers in the Kattegat, was calculated for the period 1992-09 (Table 2.3.3.2) and exploratory assessment runs were carried out. The outcome of this investigation is described in section 2.10.

Figure 2.3.3.3 illustrates the spatial distribution of 1-ringers as estimated by trawling in February 2007, 2008 and 2009. Across years, the main areas of 1-ringer distribution is in the German Bight and south of Dogger Bank, however, large catches might be seen at other sites, in 2009 such outstanding large catches were made in the Kattegat area.

The Downs herring hatch later than the autumn spawned herring and generally appears as a smaller sized group during the $1^{\text {st }}$ quarter IBTS. A recruitment index of smaller sized 1-ringers is calculated based on abundance estimates of herring $<13 \mathrm{~cm}$ (ICES CM 2000/ ACFM:12, and ICES CM 2001/ ACFM:12). Table 2.3.3.3 includes abundance estimates of 1-ringer herring smaller than 13 cm , calculated as the standard index but is in this case for herring $<13 \mathrm{~cm}$ only. Indices for these small 1-ringers are given either for the total area or the area excluding division IIIa, and their relative proportions are also shown. In the time-series, the proportion of 1-ringers smaller than 13 cm (of total catches) is in the order of $20 \%$, and the contribution from division IIIa to the overall abundance of $<13 \mathrm{~cm}$ herring varies markedly during the period (Table 2.3.3.3). About $31 \%$ of this year's group of 1-ringers is smaller than 13 cm . A large part of these are found in the IIIa (incl. Kattegat), but as mentioned above, the small 1-ringers in this area more likely stem from Western Baltic Spring Spawners.

## 2-5+ ringer abundances

Table 2.3.3.4 shows the time-series of abundance estimates of 2-5+ ringers from the $1^{\text {st }}$ quarter IBTS for the period 1983-2009. The present 2009 indices for $2-4$ ringers are very low ( $8-18 \%$ of long term means), only the index of $5+$ ringers - which includes the large 2000 year class - is of significant magnitude ( $84 \%$ of long term mean)

Table 2.3.3.1 North Sea herring. Density and abundance estimates of 0-ringers caught in February during the IBTS. Values given for year classes by areas are density estimates in numbers per square metre. Total abundance is found by multiplying density by area and summing up.

| Area | North WEST | North EAST | Central WEST | Central EAST | South WEST | South <br> EAST | Div. IIIA | SOUTH' <br> BIGHT | 0-RINGER <br> ABUNDANCE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \text { Area } \mathrm{m}^{2} \mathrm{x} \\ 10^{9} \end{gathered}$ | 83 | 34 | 86 | 102 | 37 | 93 | 31 | 31 |  |
| Year <br> class |  |  |  |  |  |  |  |  | no. in $10^{9}$ |
| 1976 | 0.054 | 0.014 | 0.122 | 0.005 | 0.008 | 0.002 | 0.002 | 0.016 | 17.1 |
| 1977 | 0.024 | 0.024 | 0.05 | 0.015 | 0.056 | 0.013 | 0.006 | 0.034 | 13.1 |
| 1978 | 0.176 | 0.031 | 0.061 | 0.02 | 0.01 | 0.005 | 0.074 | 0 | 52.1 |
| 1979 | 0.061 | 0.195 | 0.262 | 0.408 | 0.226 | 0.143 | 0.099 | 0.053 | 101.1 |
| 1980 | 0.052 | 0.001 | 0.145 | 0.115 | 0.089 | 0.339 | 0.248 | 0.187 | 76.7 |
| 1981 | 0.197 | 0 | 0.289 | 0.199 | 0.215 | 0.645 | 0.109 | 0.036 | 133.9 |
| 1982 | 0.025 | 0.011 | 0.068 | 0.248 | 0.29 | 0.309 | 0.47 | 0.14 | 91.8 |
| 1983 | 0.019 | 0.007 | 0.114 | 0.268 | 0.271 | 0.473 | 0.339 | 0.377 | 115 |
| 1984 | 0.083 | 0.019 | 0.303 | 0.259 | 0.996 | 0.718 | 0.277 | 0.298 | 181.3 |
| 1985 | 0.116 | 0.057 | 0.421 | 0.344 | 0.464 | 0.777 | 0.085 | 0.084 | 177.4 |
| 1986 | 0.317 | 0.029 | 0.73 | 0.557 | 0.83 | 0.933 | 0.048 | 0.244 | 270.9 |
| 1987 | 0.078 | 0.031 | 0.417 | 0.314 | 0.159 | 0.618 | 0.483 | 0.495 | 168.9 |
| 1988 | 0.036 | 0.02 | 0.095 | 0.096 | 0.151 | 0.411 | 0.181 | 0.016 | 71.4 |
| 1989 | 0.083 | 0.03 | 0.04 | 0.094 | 0.013 | 0.035 | 0.041 | 0 | 25.9 |
| 1990 | 0.075 | 0.053 | 0.202 | 0.158 | 0.121 | 0.198 | 0.086 | 0.196 | 69.9 |
| 1991 | 0.255 | 0.39 | 0.431 | 0.539 | 0.5 | 0.369 | 0.298 | 0.395 | 200.7 |
| 1992 | 0.168 | 0.039 | 0.672 | 0.444 | 0.734 | 0.268 | 0.345 | 0.285 | 190.1 |
| 1993 | 0.358 | 0.212 | 0.26 | 0.187 | 0.12 | 0.119 | 0.223 | 0.028 | 101.7 |
| 1994 | 0.148 | 0.024 | 0.417 | 0.381 | 0.332 | 0.148 | 0.252 | 0.169 | 126.9 |
| 1995 | 0.26 | 0.086 | 0.699 | 0.092 | 0.266 | 0.018 | 0.001 | 0.02 | 106.2 |
| 1996 | 0.003 | 0.004 | 0.935 | 0.135 | 0.436 | 0.379 | 0.039 | 0.032 | 148.1 |
| 1997 | 0.042 | 0.021 | 0.338 | 0.064 | 0.178 | 0.035 | 0.023 | 0.083 | 53.1 |
| 1998 | 0.1 | 0.056 | 1.15 | 0.592 | 0.998 | 0.265 | 0.28 | 0.127 | 244.0 |
| 1999 | 0.045 | 0.011 | 0.799 | 0.2 | 0.514 | 0.22 | 0.107 | 0.026 | 137.1 |
| 2000 | 0.284 | 0.011 | 1.052 | 0.197 | 1.156 | 0.376 | 0.063 | 0.006 | 214.8 |
| 2001 | 0.08 | 0.019 | 0.566 | 0.473 | 0.567 | 0.247 | 0.209 | 0.226 | 161.8 |
| 2002 | 0.141 | 0.04 | 0.287 | 0.028 | 0.121 | 0.045 | 0.003 | 0.157 | 54.4 |
| 2003 | 0.045 | 0.005 | 0.284 | 0.074 | 0.106 | 0.021 | 0.022 | 0.154 | 47.3 |
| 2004 | 0.017 | 0.010 | 0.189 | 0.089 | 0.268 | 0.187 | 0.027 | 0.198 | 61.3 |
| 2005 | 0.013 | 0.018 | 0.327 | 0.081 | 0.633 | 0.184 | 0.007 | 0.131 | 83.1 |
| 2006 | 0.004 | 0.001 | 0.240 | 0.025 | 0.098 | 0.018 | 0.040 | 0.228 | 37.2 |
| 2007 | 0.013 | 0.009 | 0.184 | 0.029 | 0.067 | 0.047 | 0.018 | 0.007 | 27.8 |
| 2008 | 0.145 | 0.139 | 0.277 | 0.241 | 0.101 | 0.093 | 0.160 | 0.433 | 95.8 |

Table 2.3.3.2. Test indices of 0 -ringers and 1-ringers. Used in exploration of the influence from exclusion of small larvae from Downs herring in standard 0-ringer index estimation, and of the influence from inclusion of Kattegat WBSS 1-ringers in standard 1-ringer index estimation.

| Year of sampling | Test indices of 0 -ringers | Test indices of 1-ringers |
| :---: | :---: | :---: |
| 1984 |  | 1216 |
| 1985 |  | 1868 |
| 1986 |  | 1668 |
| 1987 |  | 3514 |
| 1988 |  | 2043 |
| 1989 |  | 1730 |
| 1990 |  | 912 |
| 1991 |  | 1245 |
| 1992 | 165 | 1194 |
| 1993 | 195 | 2909 |
| 1994 | 158 | 1404 |
| 1995 | 172 | 1040 |
| 1996 | 102 | 1243 |
| 1997 | 133 | 3266 |
| 1998 | 47 | 2078 |
| 1999 | 240 | 670 |
| 2000 | 109 | 2903 |
| 2001 | 371 | 1975 |
| 2002 | 147 | 3288 |
| 2003 | 62 | 1216 |
| 2004 | 42 | 1039 |
| 2005 | 62 | 1052 |
| 2006 | 83 | 857 |
| 2007 | 49 | 1176 |
| 2008 | 40 | 1638 |
| 2009 | 93 | 1692 |

Table 2.3.3.3. North Sea herring. Indices of 1-ringers from the IBTS $1^{\text {st }}$ Quarter. Estimation of the small sized component (possibly Downs herring) in different areas. " North Sea" = total area of sampling minus IIIa.

| $\begin{aligned} & \text { YeAR } \\ & \text { CLASS } \end{aligned}$ | Year of SAMPLING | ALL <br> 1-RINGERS <br> IN TOTAL <br> AREA <br> (NO/HOUR) | Small $<13$ cm <br> 1-RINGERS <br> IN TOTAL <br> AREA <br> (NO/HOUR) | Proportion <br> OF SMALL <br> IN TOTAL <br> AREA <br> VS. ALL <br> SIZES | SmALL<13cm <br> 1-RINGERS <br> in North Sea (NO/hour) | Proportion <br> OF SMALL IN <br> North Sea <br> vs. ALL <br> SIZES | Proportion of Small in IIIA vs SMALL IN TOTAL AREA |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 1979 | 168 | 11 | 0.07 | 12 | 0.07 | 0 |
| 1978 | 1980 | 316 | 108 | 0.34 | 106 | 0.34 | 0.09 |
| 1979 | 1981 | 495 | 51 | 0.1 | 41 | 0.08 | 0.25 |
| 1980 | 1982 | 798 | 177 | 0.22 | 185 | 0.23 | 0.03 |
| 1981 | 1983 | 1270 | 192 | 0.15 | 185 | 0.15 | 0.10 |
| 1982 | 1984 | 1516 | 346 | 0.23 | 297 | 0.20 | 0.20 |
| 1983 | 1985 | 2097 | 315 | 0.15 | 298 | 0.14 | 0.12 |
| 1984 | 1986 | 2663 | 596 | 0.22 | 390 | 0.15 | 0.39 |
| 1985 | 1987 | 3693 | 628 | 0.17 | 529 | 0.14 | 0.22 |
| 1986 | 1988 | 4394 | 2371 | 0.54 | 720 | 0.16 | 0.72 |
| 1987 | 1989 | 2332 | 596 | 0.26 | 531 | 0.23 | 0.17 |
| 1988 | 1990 | 1062 | 70 | 0.07 | 62 | 0.06 | 0.18 |
| 1989 | 1991 | 1287 | 330 | 0.26 | 337 | 0.26 | 0.05 |
| 1990 | 1992 | 1268 | 125 | 0.1 | 130 | 0.10 | 0.03 |
| 1991 | 1993 | 2794 | 676 | 0.24 | 176 | 0.06 | 0.76 |
| 1992 | 1994 | 1752 | 283 | 0.16 | 240 | 0.14 | 0.21 |
| 1993 | 1995 | 1346 | 449 | 0.33 | 445 | 0.33 | 0.08 |
| 1994 | 1996 | 1891 | 604 | 0.32 | 467 | 0.25 | 0.28 |
| 1995 | 1997 | 4405 | 1356 | 0.31 | 1089 | 0.25 | 0.25 |
| 1996 | 1998 | 2276 | 1322 | 0.58 | 1399 | 0.61 | 0.02 |
| 1997 | 1999 | 753 | 152 | 0.2 | 149 | 0.20 | 0.09 |
| 1998 | 2000 | 3725 | 1117 | 0.3 | 991 | 0.27 | 0.18 |
| 1999 | 2001 | 2499 | 328 | 0.13 | 307 | 0.12 | 0.13 |
| 2000 | 2002 | 4065 | 1553 | 0.38 | 1471 | 0.36 | 0.12 |
| 2001 | 2003 | 2765 | 717 | 0.26 | 237 | 0.09 | 0.69 |
| 2002 | 2004 | 979 | 665 | 0.68 | 710 | 0.73 | 0.01 |
| 2003 | 2005 | 1002 | 340 | 0.34 | 356 | 0.36 | 0.03 |
| 2004 | 2006 | 922 | 122 | 0.13 | 128 | 0.14 | 0.02 |
| 2005 | 2007 | 1336 | 304 | 0.23 | 305 | 0.23 | 0.07 |
| 2006 | 2008 | 1901 | 440 | 0.23 | 471 | 0.25 | 0.01 |
| 2007 | 2009 | 2347 | 739 | 0.31 | 629 | 0.27 | 0.21 |

Table 2.3.3.4. North Sea herring. Indices of 2-5+ ringers from the $1^{\text {st }}$ quarter IBTS

| Year of SAMPLING | $\begin{gathered} \text { 2-RINGER } \\ \text { NO/H } \end{gathered}$ | $\begin{gathered} 3 \text {-RINGER } \\ \text { NO/H } \end{gathered}$ | $\begin{gathered} \text { 4-RINGER } \\ \text { NO/H } \\ \hline \end{gathered}$ | $\begin{gathered} 5+\text { RINGER } \\ \text { NO } / \mathrm{H} \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: |
| 1983 | 139 | 45 | 14 | 24 |
| 1984 | 161 | 61 | 27 | 10 |
| 1985 | 722 | 282 | 42 | 28 |
| 1986 | 782 | 276 | 79 | 28 |
| 1987 | 918 | 116 | 59 | 49 |
| 1988 | 4163 | 792 | 58 | 25 |
| 1989 | 875 | 339 | 89 | 9 |
| 1990 | 462 | 280 | 269 | 71 |
| 1991 | 693 | 259 | 222 | 146 |
| 1992 | 437 | 193 | 55 | 92 |
| 1993 | 787 | 223 | 45 | 66 |
| 1994 | 1167 | 213 | 69 | 43 |
| 1995 | 1393 | 279 | 37 | 7 |
| 1996 | 198 | 33 | 10 | 8 |
| 1997 | 507 | 163 | 31 | 20 |
| 1998 | 792 | 96 | 21 | 18 |
| 1999 | 451 | 501 | 98 | 36 |
| 2000 | 199 | 155 | 59 | 9 |
| 2001 | 1129 | 317 | 94 | 68 |
| 2002 | 658 | 338 | 25 | 20 |
| 2003 | 1556 | 612 | 360 | 53 |
| 2004 | 451 | 777 | 112 | 171 |
| 2005 | 214 | 356 | 389 | 131 |
| 2006 | 1464 | 330 | 252 | 339 |
| 2007 | 41 | 18 | 8 | 41 |
| 2008 | 253 | 155 | 255 | 200 |
| 2009 | 136 | 22 | 14 | 55 |



Figure 2.3.3.1. North Sea herring. Distribution of 0-ringer herring, year classes 2006-2008. Density estimates of 0-ringers within each statistical rectangle are based on MIK catches during IBTS in February 2007-2009. Areas of filled circles illustrate densities in no $\mathbf{m}^{-2}$, the area of a circle extending to the border of a rectangle represents $1 \mathrm{~m}^{-2}$


Figure 2.3.3.2 North Sea herring. Absolute ( $n 0^{*} \mathbf{1 0}^{9}$ ) and relative abundance of 0 -ringers in the area west of $2^{\circ}$ E in the North Sea. Abundances are based on MIK sampling during IBTS, the relative abundance in the western part is estimated as the number of 0 -ringers west of $2^{\circ} \mathrm{E}$ relative to total number of 0 -ringers.


Figure 2.3.3.3. North Sea herring. Distribution of 1-ringer herring, year classes 2005-2007. Density estimates of 1-ringers within each statistical rectangle are based on GOV catches during IBTS in February 2007-2009. Areas of filled circles illustrate numbers per hour, the area of a circle extending to the border of a rectangle represents $45000 \mathrm{~h}^{\mathbf{- 1}}$.

### 2.4 Mean weights-at-age and maturity-at-age

### 2.4.1 Mean weights-at-age

Table 2.4.1.1 shows the historic mean weights-at-age (winter ringers, wr) in the North Sea stock during the 3rd quarter in Divisions IV and IIIa (acoustic survey) as well as the mean weights-in-the-catch from 1996 to 2008, for comparison. The data for 2008 are taken from Table 2.3.1.2. In the third quarter most fish are approaching their peak weights just prior to spawning. The catch and acoustic survey mean weights of 1ringers in 2008 are high, but slightly lower than in 2007. The mean weights for 2008 for 2-ringers and olders are higher than in 2007 except for the 5-ringers and 7-ringers. This last class corresponds to the 2000 year class which is possibly the largest in recent years and the first large one competing with an already large herring stock biomass, grew more slowly than the other year classes. As a general rule, a decline in mean weights in the older fish (4+wr), has been observed since 1996 although the rate of decline has reduced to almost zero in recent years (Figure 2.4.1.1).

### 2.4.2 Maturity ogive

The percentages of North Sea autumn-spawning herring (at age) that are considered mature in 2008 were estimated from the acoustic survey (Table 2.4.2.1). The method and justification for the use of values derived from a single year's data was described fully in ICES (1996/ACFM:10). For 2-ringers, 3-ringers and 4-ringers, the proportions mature are $86 \%, 98 \%$ and $99 \%$, respectively (Table2.4.2.1). The percentage mature of 2 - ring fish is high but similar to 2002. For the 3 and 4-ringers, the proportions mature are high.

Mean weight for the 3-ringers in the 2007 acoustic survey (Table 2.4.1.1) is below the historical average but this has not led to a decrease in the maturity ogive. The 2000 year class, which matured more slowly, became fully mature in 2006.

Table 2.4.1.1: North Sea Herring: Mean stock weight-at-age (wr) in the third quarter, in Divisions IVa, IVb and IIIa. Mean catch weight-at-age for the same quarter and area is included for comparison. Weights-at-age in the catch for 1996 to 2001 were revised by SG Rednose, for details of the revision see the 2007 report (ICES CM 2007/ACFM:11). AS = acoustic survey.

| W. rings Year | $\begin{aligned} & \mathbf{1} \\ & \text { AS } \end{aligned}$ |  | $2$ |  | 3 AS |  | 4 |  |  |  | 6 |  |  | 3Q |  |  | 9+ ${ }^{\text {AS }}$ | $3 Q$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 45 | 75 | 119 | 135 | 196 | 186 | 253 | 224 | 262 | 229 | 299 | 253 | 306 | 292 | 325 | 300 | 335 | 302 |
| 1997 | 45 | 43 | 120 | 129 | 168 | 175 | 233 | 220 | 256 | 247 | 245 | 255 | 265 | 278 | 269 | 295 | 329 | 295 |
| 1998 | 52 | 54 | 109 | 131 | 198 | 172 | 238 | 209 | 275 | 237 | 307 | 263 | 289 | 269 | 308 | 313 | 363 | 298 |
| 1999 | 52 | 62 | 118 | 128 | 171 | 163 | 207 | 193 | 236 | 228 | 267 | 252 | 272 | 263 | 230 | 275 | 260 | 306 |
| 2000 | 46 | 54 | 118 | 123 | 180 | 172 | 218 | 201 | 232 | 228 | 261 | 241 | 295 | 266 | 300 | 286 | 280 | 271 |
| 2001 | 50 | 69 | 127 | 136 | 162 | 167 | 204 | 199 | 228 | 218 | 237 | 237 | 255 | 262 | 286 | 288 | 294 | 298 |
| 2002 | 45 | 50 | 138 | 140 | 172 | 177 | 194 | 200 | 224 | 224 | 247 | 244 | 261 | 252 | 280 | 281 | 249 | 298 |
| 2003 | 46 | 65 | 104 | 119 | 185 | 177 | 209 | 198 | 214 | 210 | 243 | 236 | 281 | 247 | 290 | 272 | 307 | 282 |
| 2004 | 35 | 45 | 116 | 125 | 139 | 159 | 206 | 203 | 231 | 234 | 253 | 250 | 262 | 264 | 279 | 262 | 270 | 299 |
| 2005 | 43 | 53 | 135 | 124 | 171 | 177 | 181 | 201 | 229 | 234 | 248 | 249 | 253 | 261 | 274 | 287 | 295 | 270 |
| 2006 | 45 | 61 | 127 | 139 | 158 | 163 | 188 | 192 | 188 | 205 | 225 | 242 | 243 | 257 | 244 | 260 | 265 | 285 |
| 2007 | 66 | 75 | 123 | 153 | 155 | 171 | 171 | 183 | 204 | 215 | 198 | 211 | 218 | 252 | 247 | 263 | 233 | 273 |
| 2008 | 62 | 67 | 141 | 151 | 180 | 192 | 183 | 207 | 194 | 211 | 230 | 240 | 217 | 243 | 268 | 276 | 282 | 312 |

Table 2.4.2.1: North Sea herring. Percentage maturity at 2, 3, 4 and 5+ ring for Autumn Spawning herring in the North Sea. The values are derived from the acoustic survey for 1988 to 2007.

| Year \ Ring | 2 | 3 | 4 | $5+$ |
| :---: | :---: | :---: | :---: | :---: |
| 1988 | 65.6 | 87.7 | 100 | 100 |
| 1989 | 78.7 | 93.9 | 100 | 100 |
| 1990 | 72.6 | 97.0 | 100 | 100 |
| 1991 | 63.8 | 98.0 | 100 | 100 |
| 1992 | 51.3 | 100 | 100 | 100 |
| 1993 | 47.1 | 62.9 | 100 | 100 |
| 1994 | 72.1 | 85.8 | 100 | 100 |
| 1995 | 72.6 | 95.4 | 100 | 100 |
| 1996 | 60.5 | 97.5 | 100 | 100 |
| 1997 | 64.0 | 94.2 | 100 | 100 |
| 1998 | 64.0 | 89.0 | 100 | 100 |
| 1999 | 81.0 | 91.0 | 100 | 100 |
| 2000 | 66.0 | 96.0 | 100 | 100 |
| 2001 | 77.0 | 92.0 | 100 | 100 |
| 2002 | 86.0 | 97.0 | 100 | 100 |
| 2003 | 43.0 | 93.0 | 100 | 100 |
| 2004 | 69.8 | 64.9 | 100 | 100 |
| 2005 | 76.0 | 97.0 | 96.0 | 100 |
| 2006 | 66.0 | 88.0 | 98.0 | 100 |
| 2007 | 71.0 | 92.0 | 93.0 | 100 |
| 2008 | 86.0 | 98.0 | 99.0 | 100 |



Figure 2.4.1.1: Figure 2.4.1.1 North Sea Herring. Mean weights for 4-ringers and older for the 3rd quarter in Divisions IV and IIIa (acoustic survey) and mean weights-in-the-catch for comparison

### 2.5 Recruitment

Information on the development in North Sea herring recruitment is available from the two IBTS indices, the 1-ringer and the 0-ringer index. Further, the ICA assessment provides estimates of the recruitment of herring in which information from the catch and from all fishery independent indices is incorporated.

### 2.5.1 Relationship between the MIK 0-ringer and the IBTS 1 -ringer indices

The 0-ringer MIK index predicts the year class strength one year before the information is available from the IBTS 1-ringer estimates. The relationship between year class estimates from the two indices is illustrated in Figure 2.5.1 and described by the fitted linear regression. Last year's prediction of the 2007 year class was much lower than indicated from this year's IBTS 1-ringer index of the year class (circled in the figure). Generally, there is a good agreement between the indices in their description of temporal trends in recruitment (Figure 2.5.2), but for the last two year classes the predicted levels of recruitment deviate. Possible explanations for this discrepancy are discussed in sections 2.3.3.1 and 2.3.3.2.

### 2.5.2 Trends in recruitment from the assessment

Abundances of recruiting North Sea herring are estimated from the assessment (see the temporal trend of recruitment in Figure 2.6.2.1). The recruitment declined during the sixties and the seventies, followed by a marked increase in the early eighties. After the strong 1985 year class recruitment declined again until the appearance of the strong year classes 1998-2001. During the following years the recruitment has generally been low. The trends in recruitment are described in detail by Payne et al. (2009). The MIK index of 0-ringer recruitment for the present year indicates a marked increase for the 2008 year class (section 2.3.3.1).


Figure 2.5.1 North Sea herring. Relationship between indices of 0 -ringers and 1-ringers for year classes 1977 to 2007. The 2007 relation is circled, the present 0 -ringer index for year class 2008 is indicated by an arrow.


Figure 2.5.2 North Sea herring. Time series of 0-ringer and 1-ringer indices. Year classes 1976 to 2008 for 0-ringers, year classes 1977-2007 for 1-ringers.

### 2.6 Assessment of North Sea herring

### 2.6.1 Data exploration and preliminary results

North Sea was classed as an update assessment in 2009 by ACOM, as a benchmark assessment took place in 2006. The choice of assessment model, catch and survey weightings and the length of separable period were not explored in 2009, and for justification of the approach refer to the benchmark assessment (HAWG 2006) and Simmonds (2003; 2009). Following the benchmark investigation in 2006, the tool for the assessment of North Sea herring is FLICA, similar to 2008.

Acoustic, bottom trawl (IBTS), MIK and larvae (MLAI) surveys are available for the assessment of North Sea autumn spawning herring. The surveys and the years for which they are available are given in Table 2.6.1.1. The WG still shares the opinion that the assessment is best executed including all surveys (Simmonds 2009).

This year's assessment is an update assessment, therefore the input data and the performance of the assessment has been carefully scrutinised to check for potential problems, but no changes to the methods or development of the model took place in 2009. The diagnostics do not indicate any significant pattern or unreliable data points (Figure 2.6.1.1 to Figure 2.6.1.16). There is no evidence of cohort effects during the full selection pattern of the separable period. Overall the catch residuals are small. There has been no major change in the patterns in the residuals for the surveys by adding the extra year of data and the contribution of each indices to the objective function is in range with the expected. (Figures 2.6.1.17 and 2.6.1.18).

It has been noted that in recent years the MLAI has positive residuals (Figure 2.6.1.17) and the acoustic survey has a block of negative residuals at older ages. The current assessment shows that this pattern for the MLAI and acoustics has been maintained (Figure 2.6.1.17). However, in 2006 the residual from the MLAI is small. In the 2006 benchmark assessment it was concluded that one of the reasons for the relatively stable assessment was the balance of the major sources of information, with each potentially delivering short periods with bias but in combination providing a balance of errors.

Figures 2.6.1.29 to 2.6.1.21 shows retrospective estimates of SSB , recruitment, mean $\mathrm{F}_{2}$ ${ }^{6}$, selectivity pattern and year class cohorts, by removing one year of data at a time, up to 8 years in total. The estimation of $F$ shows considerable consistency over the last 10 years. SSB is reasonable consistent over the last 6 years. The retrospective estimates for recruitment in the years 2003 to 2007 deviate from each other in a larger extend than observed within the SSB and $\mathrm{F}_{2-6}$ retrospectives, for the first year of estimation. However, for all three historical retrospective estimates it can be concluded that the patterns are consistent. Figure 2.6.1.21 shows the retrospective pattern of the number per cohort. This pattern is consistent over the years as well. The selectivity pattern has not changed greatly over the recent period, however selectivity increased for the 5 to 7 year olds for 4 years in a row (Figure 2.6.1.19). Figure 2.6.1.22 shows the 'otolith' plot, representing the uncertainty of the fit of the assessment model. The $90 \%$ confidence interval of SSB indicates that the stock is above $\mathrm{Blim}_{\mathrm{lim}}$.

Further data screening of the input data on mature - immature biomass ratio's, survey CPUE's, proportion of catch numbers and weights at age and proportion of IBTS and acoustic survey ages have been executed, as well as correlation coefficient analy-
ses for the acoustic and IBTS survey (see Figures 2.6.1.23 to 2.6.1.30). No issues were raised by this exercise.

### 2.6.2 Final Assessment for NS herring

In according with the settings described in the stock annex, the final assessment of North Sea herring was carried out by fitting the integrated catch-at-age model (ICA, in the FLR environment - version 1.4-9 - Sat Mar 21 13:07:02 2009)). The input data and model settings are shown in Table 2.6.2.1 - Table 2.6.2.11, the ICA output is presented in Table 2.6.2.12 - Table 2.6.2.21, the stock summary in Table 2.6.2.12 and Figure 2.6.2.1 and model fit and parameter estimates in Table 2.6.2.21. Diagnostics of the catch for the separable period are shown in Figure 2.6.2.2. In Figure 2.6.2.3 Yield to F and Yield to SSB curves are shown, also indicating reference points $B_{\text {msy }}$ and $F_{\text {msy. }}$ Figure 2.6.2.4 shows the agreed management plan including the biomass trigger points and contains the $\mathrm{F}_{2-6}$ estimates of the past 7 years.

The spawning stock at spawning time in 2008 is estimated at approximately 1.0 million tonnes, increasing from 0.95 million tonnes in 2007. The estimate of 0 -wr fish in 2009 (2008 year class) is 1.6 times bigger than in 2008 and $81 \%$ of the geometric mean of recruitment since 1981 (see Table 2.6.2.15). This new recruitment is higher than the previous 6 years but similar to the 2001 year class (Figure 2.6.2.5). It does not suggest that recruitment has recovered. The strong 2000 year class is still evident in the population, at $7-\mathrm{wr}$ in 2008. Mean $\mathrm{F}_{2-6}$ in 2008 is estimated at approximately 0.24 , which is above the management agreement target F , while mean $\mathrm{F}_{0-1}$ is 0.05 , approximately at the agreed target, and lower than 2007.

### 2.6.3 State of the Stock

| Spawning <br> biomass in <br> relation to pre- <br> cautionary <br> limits | Fishing mor- <br> tality in rela- <br> tion to <br> precautionary <br> limits | Fishing mor- <br> tality in rela- <br> tion to highest <br> yield | Fishing mor- <br> tality in rela- <br> tion to <br> agreed target | Comment |
| :--- | :--- | :--- | :--- | :--- |

Based on the most recent estimates of SSB and fishing mortality, ICES classifies the stock as being at risk of having reduced reproductive capacity and harvested sustainably. The SSB in autumn 2008 was estimated at 1.0 million $t$, and is expected to remain below Bpa ( 1.3 million $t$ ) in 2009. $\mathrm{F}_{2-6}$ in 2008 was estimated at 0.24 , above the target $\mathrm{F}_{2-6}$ of 0.14 . The year classes since 2002 are estimated to be among the weakest since the late 1970s.

Table 2.6.1.1 North Sea herring. Years of duration of survey and years used in the assessment.

| SURVEY |  | YEARS SURVEY HAS | YEARS USED IN |
| :--- | :--- | :--- | :--- |
|  | AGE RANGE | BEEN RUNNING | ASSESSMENT |
| MLAI (Larvae survey) | SSB | $1972-2008$ | $1973-2008$ |
| IBTS 1 ${ }^{\text {st }}$ Quarter (Trawl survey) | $1-5 w r$ | $1971-2009$ | $1984-2009$ |
| Acoustic (+trawl) | $1 w r$ | $1995-2008$ | $1997-2008$ |
|  | $2-9+w r$ | $1984-2008$ | $1989-2008$ |
| MIK net | $0 w r$ | $1977-2009$ | $1992-2009$ |

## TABLE 2.6.2.1 North Sea Autumn Spawning Herring. Catch in Number

| year Units |  |  | thousands |  | 1964 | 1965 | -1966 | 1967 | 71968 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1960 | 1961 | 1962 | 1963 |  |  |  |  |  |  |
| 0 | 194600 | 1269200 | 141800 | 442800 | 496900 | 157100 | 374500 | 645400 | - 8393 | 00 |
| 1 | 2392700 | 336000 | 2146900 | 1262200 | 2971700 | 3209300 | 1383100 | 1674300 | 024250 |  |
| 2 | 1142300 | 1889400 | 269600 | 2961200 | 1547500 | 2217600 | 2569700 | 1171500 | 017952 |  |
| 3 | 1966700 | 479900 | 797400 | 177200 | 2243100 | 1324600 | 741200 | 1364700 | 014943 |  |
| 4 | 165900 | 1455900 | 335100 | 158300 | 148400 | 2039400 | 450100 | 371500 | 06214 | 00 |
| 5 | 167700 | 124000 | 1081800 | 80600 | 149000 | 145100 | 889800 | 297800 | 01571 |  |
| 6 | 112900 | 157900 | 126900 | 229700 | - 95000 | 151900 | - 45300 | 393100 | -14500 |  |
| 7 | 125800 | 61400 | 145100 | 22400 | 256300 | 117600 | - 64800 | 67900 | 01634 |  |
| 8 | 128600 | 56000 | - 86300 | 42000 | - 26300 | 413000 | 95500 | 81600 | 0137 | 00 |
| 9 | 142000 | 87500 | - 86800 | 51000 | - 57700 | 78400 | - 236300 | 172800 |  |  |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 |
| 0 | 112000 | 898100 | 684000 | 750400 | 289400 | 996100 | 263800 | 238200 | 256800 | 130000 |
| 1 | 2503300 | 1196200 | 4378500 | 3340600 | 2368000 | 846100 | 2460500 | 126600 | 144300 | 168600 |
| 2 | 1883000 | 2002800 | 1146800 | 1440500 | 1344200 | 772600 | 541700 | 901500 | 44700 | 4900 |
| 3 | 296300 | 883600 | - 662500 | 343800 | - 659200 | 362000 | 259600 | 117300 | 186400 | 5700 |
| 4 | 133100 | 125200 | 208300 | 130600 | 150200 | 126000 | 140500 | 52000 | 10800 | 5000 |
| 5 | 190800 | 50300 | 26900 | 32900 | - 59300 | 56100 | 57200 | 34500 | 7000 | 300 |
| 6 | 49900 | 61000 | 30500 | 5000 | 30600 | 22300 | 16100 | 6100 | 4100 | 200 |
| 7 | 42700 | 7900 | 26800 | 200 | 3700 | 5000 | 9100 | 4400 | 1500 | 200 |
| 8 | 27400 | 12000 | 100 | 1100 | - 1400 | 2000 | 3400 | 1000 | 700 | 200 |
| 9 | 25100 | 12200 | 12400 | 400 | -600 | 1100 | 1400 | 400 | 0 | 300 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1979 | 1980 | 1981 | 1982 | 1983 | 198 | 1985 | 198 | 86 | 1987 |
| $\bigcirc$ | 542000 | 1262700 | 9519700 | 11956700 | 13296900 | 697330 | 4211000 | 372470 | 008229 | 200 |
| 1 | 159200 | 245100 | 872000 | 1116400 | 2448600 | 181840 | 3253000 | - 480140 | 0068363 | 300 |
| 2 | 34100 | 134000 | 284300 | 299400 | - 573800 | - 114620 | 1326300 | 126670 | 002137 | 200 |
| 3 | 10000 | 91800 | 56900 | 230100 | - 216400 | - 44140 | 00 1182400 | - 84080 |  | 900 |
| 4 | 10100 | 32200 | 39500 | 33700 | 105100 | - 20150 | 00 368500 | - 46590 | 004671 | 100 |
| 5 | 2100 | 21700 | 28500 | 14400 | 26200 | -8110 | 124500 | 12980 |  | 800 |
| 6 | 200 | 2300 | 22700 | 6800 | - 22800 | 2260 | 43600 | 06210 |  | 4700 |
| 7 | 800 | 1400 | 18700 | 7800 | 12800 | - 2520 | 20200 | 02050 |  | 3800 |
| 8 | 600 | 400 | 5500 | 3600 | 11000 | -1110 | 13100 | - 1360 |  | 8000 |
| 9 | 100 | 100 | 1100 | 1100 | 12100 | 01860 | 0016100 | 01480 |  | 200 |
| year 100 12100 188000 |  |  |  |  |  |  |  |  |  |  |
| age | 1988 | 1989 | 91990 | 1991 | 1992 |  | 1993 | 9419 | 995 | 1996 |
| 0 | 3164800 | 3057800 | 1302800 | 2386600 | 10331300 | 102654 | 4004498900 | 0074384 | 469231 | 11226 |
| 1 | 7867000 | 3145900 | 3020000 | 2138900 | 2303100 | - 38268 | 1785200 | 0016648 | 874160 | 06393 |
| 2 | 2232500 | 1593700 | 899300 | 1132800 | 1284900 | 11763 | 3001783200 | 200 14440 | 0616 | 42084 |
| 3 | 1090700 | 1363800 | 779100 | 556700 | - 442700 | 06090 | 48900 4800 | 008167 | 70352 | 25601 |
| 4 | 383700 | 809300 | - 861000 | 548900 | 361500 | - 3055 | 347600 | 02317 | 79417 | 2099 |
| 5 | 255800 | 211800 | - 387500 | 501200 | - 360500 | 02156 | 600109000 | 1185 | 536 | 57586 |
| 6 | 128100 | 123700 | 80200 | 205300 | - 375600 | 02260 | 00091800 | 00 551 | 128 | 22534 |
| 7 | 38000 | 61000 | - 54400 | 39300 | 152400 | -1880 | 7600 7600 | 00414 | 409 | 9264 |
| 8 | 15300 | 19500 | 28800 | 25600 | - 39200 | - 873 | 30070000 | 00689 | 955 | 17195 |
| 9 | 8500 | 8700 | - 11900 | 13000 | 23300 | - 4170 | 70046600 | 022 | 245 | 3948 |
| age | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |  |
| $\bigcirc$ | 431175 | 259526 | 1566349 | 1105085 | 18326917 | 730279 | 369074 | 715597 | 101555 |  |
| 1 | 479702 | 977680 | 303520 | 1171677 | 614469 | 837557 | 617021 | 206648 | 715547 |  |
| 2 | 687920 | 1220105 | 616354 | 622853 | 8426355 | 5795921 | 1221992 | 447918 | 355453 |  |
| 3 | 446909 | 537932 | 1058716 | 463170 | 4856289 | 970577 | 52938613 | 366155 | 485746 |  |
| 4 | 284920 | 276333 | 294066 | 646814 | 278884 | 292205 | 835552 | 543376 | 1318647 |  |
| 5 | 109178 | 175817 | 135648 | 213466 | 321743 | 140701 | 244780 | 753231 | 479961 |  |
| 6 | 31389 | 88927 | 69299 | 82481 | 909181 | 174570 | 107751 | 169324 | 57615 |  |
| 7 | 11832 | 15232 | 27998 | 35706 | 38252 | 48908 | 123291 | 104945 | 115212 |  |
| 8 | 18770 | 16766 | 10174 | 14624 | 17910 | 34620 | 37671 | 65341 | 88311 |  |
| 9 | 5697 | 3784 | 2054 | 2463 | 2692 | 8702 | 9044 | 31801 | 58497 |  |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2006 | 2007 | 2008 |  |  |  |  |  |  |  |
| $\bigcirc$ | 878637 | 6210057 | 798284 |  |  |  |  |  |  |  |
| 1 | 222111 | 2355532 | 235022 |  |  |  |  |  |  |  |
| 2 | 401087 | 219115 | 331772 |  |  |  |  |  |  |  |
| 3 | 310602 | 4174521 | 184771 |  |  |  |  |  |  |  |
| 4 | 464620 | 2857461 | 199069 |  |  |  |  |  |  |  |
| 5 | 997782 | 3094541 | 137529 |  |  |  |  |  |  |  |
| 6 | 252150 | 6291871 | 118349 |  |  |  |  |  |  |  |
| 7 | 247042 | 1478302 | 215542 |  |  |  |  |  |  |  |
| 8 | 63035 | 133388 | 74339 |  |  |  |  |  |  |  |
| 9 | 43377 | 23362 | 42919 |  |  |  |  |  |  |  |

Table 2.6.2.2 North Sea Autumn Spawning Herring. Weights at age in the Catch

```
Units : kg
    year
age 1960
    0 0.015 0.015 0.015 0.015 0.015 0.015 0.015 0.015 0.015 0.015 0.015 0.015
    1 0.050 0.050 0.050 0.050 0.050 0.050 0.050 0.050 0.050 0.050 0.050 0.050
    2 0.126 0.126 0.126 0.126 0.126 0.126 0.126 0.126 0.126 0.126 0.126 0.126
    3 0.176 0.176 0.176 0.176 0.176 0.176 0.176 0.176 0.176 0.176 0.176 0.176
    4 0.211 0.211 0.211 0.211 0.211 0.211 0.211 0.211 0.211 0.211 0.211 0.211
    5 0.243 0.243 0.243 0.243 0.243 0.243 0.243 0.243 0.243 0.243 0.243 0.243
    6 0.251 0.251 0.251 0.251 0.251 0.251 0.251 0.251 0.251 0.251 0.251 0.251
    7 0.267 0.267 0.267 0.267 0.267 0.267 0.267 0.267 0.267 0.267 0.267 0. 267
    8 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271
    9 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271
        year
age 1972 1973 1974 1975 1976 1977 1978 1979 1980
    0 0.015 0.015 0.015 0.015 0.015 0.015 0.015 0.015 0.015 0.007 0.010 0.010
    1 0.050 0.050 0.050 0.050 0.050 0.050 0.050 0.050 0.050 0.049 0.059 0.059
    2 0.126 0.126 0.126 0.126 0.126 0.126 0.126 0.126 0.126 0.118 0.118 0.118
    3 0.176 0.176 0.176 0.176 0.176 0.176 0.176 0.176 0.176 0.142 0.149 0.149
    4 0.211 0.211 0.211 0.211 0.211 0.211 0.211 0.211 0.211 0.189 0.179 0.179
    5 0.243 0.243 0.243 0.243 0.243 0.243 0.243 0.243 0.243 0.211 0.217 0.217
    6 0.251 0.251 0.251 0.251 0.251 0.251 0.251 0.251 0.251 0.222 0.238 0.238
    7 0.267 0.267 0.267 0.267 0.267 0.267 0.267 0.267 0.267 0.267 0.265 0.265
    8 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.271 0.274 0.274
    9 0.271 0.271 0.271 0.271 0.271 0.000 0.271 0.271 0.271 0.271 0.275 0.275
        year
```



```
    0 0.010 0.009 0.006 0.011 0.011 0.017 0.019 0.017 0.010 0.010 0.006 0.009
    1 0.059 0.036 0.067 0.035 0.055 0.043 0.055 0.058 0.053 0.033 0.056 0.042
    2 0.118 0.128 0.121 0.099 0.111 0.115 0.114 0.130 0.102 0.115 0.130 0.130
    3 0.149 0.164 0.153 0.150 0.145 0.153 0.149 0.166 0.175 0.145 0.159 0.169
    4 0.179 0.194 0.182 0.180 0.174 0.173 0.177 0.184 0.189 0.189 0.181 0.198
    5 0.217 0.211 0.208 0.211 0.197 0.208 0.193 0.203 0.207 0.204 0.214 0.207
    6 0.238 0.220 0.221 0.234 0.216 0.231 0.229 0.217 0.223 0.228 0.240 0.243
    7 0.265 0.258 0.238 0.258 0.237 0.247 0.236 0.235 0.237 0.244 0.255 0.247
    8 0.274 0.270 0.252 0.277 0.253 0.265 0.250 0.259 0.249 0.256 0.273 0.283
    9 0.275 0.292 0.262 0.299 0.263 0.259 0.287 0.271 0.287 0.310 0.281 0.276
        year
age 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 
    0 0.015 0.015 0.021 0.009 0.015 0.012 0.012 0.014 0.014 0.011 0.010 0.0124
    1 0.018 0.044 0.051 0.045 0.033 0.048 0.037 0.037 0.036 0.044 0.049 0.0638
    2 0.112 0.108 0.114 0.115 0.113 0.118 0.118 0.104 0.100 0.099 0.117 0.1214
    3 0.156 0.148 0.145 0.151 0.157 0.149 0.153 0.158 0.138 0.153 0.144 0.1513
    4 0.188 0.195 0.183 0.171 0.179 0.177 0.170 0.174 0.183 0.166 0.172 0.1634
    5 0.204 0.227 0.219 0.207 0.201 0.198 0.199 0.184 0.201 0.208 0.181 0.1933
    6 0.212 0.226 0.238 0.233 0.216 0.213 0.214 0.205 0.216 0.223 0.220 0.1900
    7 0.261 0.235 0.247 0.245 0.246 0.238 0.228 0.222 0.228 0.240 0.237 0.2232
    8 0.280 0.244 0.289 0.261 0.275 0.267 0.250 0.232 0.246 0.257 0.235 0. 2349
    0 0.288 0.291 0.283 0.301 0.262 0.288 0.252 0.256 0.272 0.278 0.262 0. 2523
        year
age 2008
    0 0.0079
    1 0.0535
    2 0.1288
    3 0.1796
    4 0.1812
    50.1832
    6 0.2157
    7 0.2161
    80.2560
    9 0.2726
```


## Table 2.6.2.3 North Sea Autumn Spawning Herring. Weights at age in the Stock.

Units : kg
year
age $\begin{array}{lllllllllllll}1960 & 1961 & 1962 & 1963 & 1964 & 1965 & 1966 & 1967 & 1968 & 1969 & 1970 & 1971\end{array}$
00.0150 .0150 .0150 .0150 .0150 .0150 .0150 .0150 .0150 .0150 .0150 .015
10.0500 .0500 .0500 .0500 .0500 .0500 .0500 .0500 .0500 .0500 .0500 .050
20.1550 .1550 .1550 .1550 .1550 .1550 .1550 .1550 .1550 .1550 .1550 .155
30.1870 .1870 .1870 .1870 .1870 .1870 .1870 .1870 .1870 .1870 .1870 .187
40.2230 .2230 .2230 .2230 .2230 .2230 .2230 .2230 .2230 .2230 .2230 .223
50.2390 .2390 .2390 .2390 .2390 .2390 .2390 .2390 .2390 .2390 .2390 .239
60.2760 .2760 .2760 .2760 .2760 .2760 .2760 .2760 .2760 .2760 .2760 .276
70.2990 .2990 .2990 .2990 .2990 .2990 .2990 .2990 .2990 .2990 .2990 .299
80.3060 .3060 .3060 .3060 .3060 .3060 .3060 .3060 .3060 .3060 .3060 .306
90.3120 .3120 .3120 .3120 .3120 .3120 .3120 .3120 .3120 .3120 .3120 .312 year
age $197219731974 \begin{array}{llllllllll}1975 & 1976 & 1977 & 1978 & 1979 & 1980 & 1981 & 1982 & 1983\end{array}$
00.0150 .0150 .0150 .0150 .0150 .0150 .0150 .0150 .0150 .0150 .0150 .017
10.0500 .0500 .0500 .0500 .0500 .0500 .0500 .0500 .0500 .0500 .0500 .057
20.1550 .1550 .1550 .1550 .1550 .1550 .1550 .1550 .1550 .1550 .1550 .150
$30.1870 .1870 .1870 .187 \quad 0.1870 .1870 .187 \quad 0.1870 .1870 .1870 .1870 .190$
40.2230 .2230 .2230 .2230 .2230 .2230 .2230 .2230 .2230 .2230 .2230 .230
50.2390 .2390 .2390 .2390 .2390 .2390 .2390 .2390 .2390 .2390 .2390 .243
60.2760 .2760 .2760 .2760 .2760 .2760 .2760 .2760 .2760 .2760 .2760 .282
70.2990 .2990 .2990 .2990 .2990 .2990 .2990 .2990 .2990 .2990 .2990 .311
80.3060 .3060 .3060 .3060 .3060 .3060 .3060 .3060 .3060 .3060 .3060 .338
90.3120 .3120 .3120 .3120 .3120 .3120 .3120 .3120 .3120 .3120 .3120 .347 year
age $191984 \quad 1985 \quad 198619871988$
00.0160 .0140 .0090 .008 0.009 0.012 0.011 0.010 0.006 0.007 0.006 0.006
10.0560 .0610 .0500 .0480 .0440 .0520 .0590 .0640 .0610 .0600 .0570 .054
20.1380 .1300 .1220 .1230 .1220 .1260 .1390 .1370 .1340 .1260 .1290 .130
30.1870 .1830 .1700 .1660 .1650 .1740 .1840 .1940 .1840 .1920 .1860 .199
40.2320 .2320 .2120 .2080 .2050 .2120 .2120 .2140 .2130 .2140 .2110 .227
50.2470 .2520 .2300 .2290 .2280 .2440 .2390 .2340 .2340 .2400 .2240 .234
60.2750 .2730 .2420 .2480 .2520 .2710 .2650 .2530 .2620 .2750 .2680 .274
70.3210 .3150 .2750 .2590 .2610 .2840 .2800 .2720 .2730 .2910 .2930 .301
80.3410 .3310 .2680 .2630 .2770 .2980 .3000 .2910 .3020 .3090 .3180 .323
90.3650 .3920 .3430 .3250 .3150 .3310 .3280 .3120 .3200 .3370 .3450 .343 year
age $199619971998 \quad 1999 \quad 2000 \quad 2001 \quad 2002 \quad 2003 \quad 2004 \quad 2005 \quad 2006 \quad 2007$ 00.0050 .0060 .0060 .0060 .0060 .0060 .0070 .0070 .0060 .0070 .0060 .008
10.0490 .0470 .0510 .0510 .0510 .0470 .047 0.042 0.041 0.041 0.051 0.055
20.1230 .1160 .1160 .1160 .1220 .1280 .1230 .1190 .1180 .1260 .1280 .125
30.1830 .1870 .1790 .1840 .1720 .1720 .1730 .1650 .1640 .1550 .1610 .156
40.2300 .2410 .2260 .2210 .2100 .2050 .2020 .2030 .1980 .1910 .1800 .180
50.2370 .2640 .2560 .2480 .2330 .2280 .2220 .2230 .2250 .2160 .2070 .196
60.2570 .2840 .2730 .2790 .2550 .2480 .2420 .2480 .2480 .2420 .2240 .212
70.2800 .2870 .2760 .2860 .2750 .2700 .2660 .2680 .2650 .2520 .2380 .230
80.3030 .3010 .2700 .2810 .2740 .2890 .2850 .2830 .2810 .2660 .2550 .245
90.3340 .3420 .3180 .3030 .2800 .2750 .2830 .2750 .2910 .2770 .2640 .249 year
age 2008
00.008
10.058
20.130
30.164
40.181
50.195
60.218
70.226
80.253
90.260

Table 2.6.2.4 North Sea Autumn Spawning Herring. Natural Mortality.

| Uni | s | NA |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 |
| 0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| 1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| 2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |
| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 5 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 7 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 8 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 9 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| 1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| 2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |
| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 5 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 7 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 8 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 9 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| 1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| 2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |
| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 5 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 7 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 8 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 9 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 2005 | 2006 | 2007 | 2008 |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 1.0 | 1.0 | 1.0 | 1.0 |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1.0 | 1.0 | 1.0 | 1.0 |  |  |  |  |  |  |  |  |  |  |  |
| 2 | 0.3 | 0.3 | 0.3 | 0.3 |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.2 | 0.2 | 0.2 | 0.2 |  |  |  |  |  |  |  |  |  |  |  |
| 4 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |  |  |  |
| 6 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |  |  |  |
| 7 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |  |  |  |
| 8 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |  |  |  |
| 9 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |  |  |  |

Table 2.6.2.5 North Sea Autumn Spawning Herring. Proportion Mature

| Units : year | NA |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| 21 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 0.82 | 0.82 | 0.82 |
| 31 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1.00 | 1.00 | 1.00 |
| 41 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1.00 | 1.00 | 1.00 |
| 51 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1.00 | 1.00 | 1.00 |
| 61 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1.00 | 1.00 | 1.00 |
| 71 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1.00 | 1.00 | 1.00 |
| 81 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1.00 | 1.00 | 1.00 |
| 91 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1.00 | 1.00 | 1.00 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 00.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 | 0.00 | 0.0 | 0.00 | 0.00 |
| 10.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 | 0.00 | 0.0 | 0.00 | 0.00 |
| 20.82 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 | 0.7 | 0.75 | 0.8 | 0.85 | 0.82 |
| 31.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.0 | 1.00 | 1.0 | 0.93 | 0.94 |
| 41.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.0 | 1.00 | 1.0 | 1.00 | 1.00 |
| 51.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.0 | 1.00 | 1.0 | 1.00 | 1.00 |
| 61.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.0 | 1.00 | 1.0 | 1.00 | 1.00 |
| 71.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.0 | 1.00 | 1.0 | 1.00 | 1.00 |
| 81.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.0 | 1.00 | 1.0 | 1.00 | 1.00 |
| 91.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.0 | 1.00 | 1.0 | 1.00 | 1.00 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 00.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 10.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 20.91 | 0.86 | 0.50 | 0.47 | 0.73 | 0.67 | 0.61 | 0.64 | 0.64 | 0.69 | 0.67 | 0.77 | 0.87 | 0.43 | 0.70 |
| 30.97 | 0.99 | 0.99 | 0.61 | 0.93 | 0.95 | 0.98 | 0.94 | 0.89 | 0.91 | 0.96 | 0.92 | 0.97 | 0.93 | 0.65 |
| 41.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 51.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 61.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 71.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 81.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 91.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age 2005 | 2006 | 2007 | 2008 |  |  |  |  |  |  |  |  |  |  |  |
| 00.00 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |
| 10.00 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |
| 20.76 | 0.66 | 0.71 | 0.86 |  |  |  |  |  |  |  |  |  |  |  |
| 30.96 | 0.88 | 0.92 | 0.98 |  |  |  |  |  |  |  |  |  |  |  |
| 40.96 | 0.98 | 0.93 | 0.99 |  |  |  |  |  |  |  |  |  |  |  |
| 51.00 | 1.00 | 1.00 | 1.00 |  |  |  |  |  |  |  |  |  |  |  |
| 61.00 | 1.00 | 1.00 | 1.00 |  |  |  |  |  |  |  |  |  |  |  |
| 71.00 | 1.00 | 1.00 | 1.00 |  |  |  |  |  |  |  |  |  |  |  |
| 81.00 | 1.00 | 1.00 | 1.00 |  |  |  |  |  |  |  |  |  |  |  |
| 91.00 | 1.00 | 1.00 | 1.00 |  |  |  |  |  |  |  |  |  |  |  |

Table 2.6.2.6 North Sea Autumn Spawning Herring. Fraction of Harvest before Spawning

```
Units : NA
    year
age 1960 1961 1962 1963 1964 1965 1966 1967 1968 1969 1970 1971 1972 1973 1974
    0 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
        year
age 1975 1976 1977 1978 1979 1980 1981 1982 1983 1984 1985 1986 1987 1988 1989
    0 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    year
age 1990 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004
    0 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    year
age 2005 2006 2007 2008
    0 0.67 0.67 0.67 0.67
    1 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67
    80.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67
```

Table 2.6.2.7 North Sea Autumn Spawning Herring. Fraction of Natural Mortality before Spawning

```
Units : NA
    year
age 1960 1961 1962 1963 1964 1965 1966 1967 1968 1969 1970 1971 1972 1973 1974
    0 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    year
age 1975 1976 1977 1978 1979 1980 1981 1982 1983 1984 1985 1986 1987 1988 1989
    0 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
        year
age 1990 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004
    0 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
        year
age 2005 2006 2007 2008
    0 0.67 0.67 0.67 0.67
    1 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67
```

Table 2.6.2.8 North Sea Autumn Spawning Herring. Survey Indicies

```
MLAI - Configuration
"Herring" "in" "Sub-area" "IV," "Divisions" "VIId" "&" "IIIa" "(autumn-
spawners)"
min max plusgroup minyear maxyear startf endf
Index type : biomass
MLAI - Index Values
Units : NA
age 
    all 13.182 7.943 2.819 2.494 6.151 7.427 14.363 9.771 14.337 20.891 26.804
        year
```



```
    all 48.367 73.818 38.444 67.69 134.382 131.732 171.592 90.332 42.147 30.069
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 
    all 20.798 22.353 43.982 56.462 72.912 60.531 40.441 129.562 109.899 267.813
        year
age 2004 2005 2006 2007 2008
    all 321.658 192.266 117.856 173.003 181.746
MLAI - Index Variance (Inverse Weights)
Units : NA
            year
age 1973 1974 1975 1976 1977 1978 1979 1980
    all 1.666667 1.666667 1.666667 1.666667 1.666667 1.666667 1.666667 1.666667
        year
```



```
    all 1.666667 1.666667 1.666667 1.666667 1.666667 1.666667 1.666667 1.666667
        year 1989 1990 1991 1992 1993 1994 1995 109 
```



```
        year
    age 1907 1998 1999 2000 2001 2002 2003 2004
    all 1.666667 1.666667 1.666667 1.666667 1.666667 1.666667 1.666667 1.666667
        year
age
```



```
MIK 0-wr - Configuration
"Herring in Sub-area IV, Divisions VIId & IIIa (autumn-spawners) . Imported from VPA
file."
min max plusgroup minyear maxyear startf endf
Index type : number
MIK 0-wr - Index Values
Units : NA
    year
age }199
    0 200.7 190.1 101.7 127 106.5 148.1 53.1 244 137.1 214.8 161.8 54.4 47.3
    year
age 2005 2006 2007 2008 2009
    0 61.3 83.1 37.2 27.8 95.8
MIK 0-wr - Index Variance (Inverse Weights)
Units : NA
    year
age 
    0 1.587302 1.587302 1.587302 1.587302 1.587302 1.587302 1.587302 1.587302
    year
age 2000 2001 2002 2003 2004 2005 2006 2007
    0 1.587302 1.587302 1.587302 1.587302 1.587302 1.587302 1.587302 1.587302
    year
age 2008 2009
    0 1.587302 1.587302
IBTS1: 1-5+ wr - Configuration
"Herring in Sub-area IV, Divisions VIId & IIIa (autumn-spawners) . Imported from VPA
file."
\begin{tabular}{rrrrrrr} 
min & max & plusgroup & minyear & maxyear & startf & endf \\
1.00 & 5.00 & 5.00 & 1984.00 & 2009.00 & 0.08 & 0.17
\end{tabular}
Index type : number
```

| Units : year |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 1 | 1515.627 | 2097.280 | 2662.812 | 3692.965 | 4394.168 | 2331.566 | 1061.572 | 1286.747 |
| 2 | 161.480 | 721.646 | 782.122 | 917.550 | 4163.384 | 875.336 | 462.097 | 693.020 |
| 3 | 61.428 | 281.990 | 276.031 | 116.315 | 791.528 | 338.514 | 279.780 | 258.604 |
| 4 | 26.888 | 42.088 | 79.007 | 59.351 | 57.957 | 89.381 | 269.108 | 221.523 |
| 5 | 10.238 | 27.941 | 28.076 | 48.763 | 25.054 | 8.519 | 71.303 | 146.096 |
| year |  |  |  |  |  |  |  |  |
| age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 1 | 1268.145 | 2794.007 | 1752.053 | 1345.754 | 1890.872 | 4404.647 | 2275.845 | 752.862 |
| 2 | 436.563 | 787.421 | 1167. 221 | 1392.857 | 197.522 | 506.536 | 791.593 | 450.623 |
| 3 | 193.085 | 222.585 | 213.059 | 278.544 | 32.875 | 162.660 | 95.660 | 501.325 |
| 4 | 54.810 | 45.042 | 69.004 | 36.670 | 10.193 | 30.532 | 20.810 | 98.179 |
| 5 | 92.268 | 65.534 | 42.503 | 6.551 | 8.079 | 19.935 | 17.841 | 35.566 |
| year |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 1 | 3725.131 | 2499.391 | 4064.829 | 2765.059 | 979.101 | 1001.585 | 911.241 | 1321.005 |
| 2 | 199.374 | 1129.308 | 658.154 | 1556.082 | 436.519 | 214.191 | 1481.330 | 50.003 |
| 3 | 154.691 | 317.069 | 338.153 | 611.890 | 766.031 | 356.007 | 334.732 | 18.250 |
| 4 | 58.838 | 93.886 | 25.048 | 359.989 | 112.374 | 388.922 | 241.137 | 7.937 |
| 5 | 8.952 | 68.284 | 19.936 | 53.166 | 170.998 | 131.481 | 328.414 | 41.284 |
| year |  |  |  |  |  |  |  |  |
| age | 2008 | 2009 |  |  |  |  |  |  |
| 1 | 1901.333 | 2346.830 |  |  |  |  |  |  |
| 2 | 252.644 | 135.824 |  |  |  |  |  |  |
| 3 | 155.190 | 22.304 |  |  |  |  |  |  |
| 4 | 254.853 | 13.528 |  |  |  |  |  |  |
| 5 | 200.270 | 55.003 |  |  |  |  |  |  |

IBTS1: 1-5+ wr - Index Variance (Inverse Weights)

|  | $\begin{aligned} & \text { ts : NA } \\ & \text { year } \end{aligned}$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 1 | 2.127660 | 2.127660 | 2.127660 | 2.127660 | 2.127660 | 2.127660 |
| 2 | 3.571429 | 3.571429 | 3.571429 | 3.571429 | 3.571429 | 3.571429 |
| 3 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
| 4 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
| 5 | $100.000000$ year | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 1 | 2.127660 | 2.127660 | 2.127660 | 2.127660 | 2.127660 | 2.127660 |
| 2 | 3.571429 | 3.571429 | 3.571429 | 3.571429 | 3.571429 | 3.571429 |
| 3 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
| 4 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
|  | $\begin{aligned} & 100.000000 \\ & \text { year } \end{aligned}$ | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
| age | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | 2.127660 | 2.127660 | 2.127660 | 2.127660 | 2.127660 | 2.127660 |
| 2 | 3.571429 | 3.571429 | 3.571429 | 3.571429 | 3.571429 | 3.571429 |
| 3 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
| 4 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
|  | $100.000000$ year | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
| age | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 1 | 2.127660 | 2.127660 | 2.127660 | 2.127660 | 2.127660 | 2.127660 |
| 2 | 3.571429 | 3.571429 | 3.571429 | 3.571429 | 3.571429 | 3.571429 |
| 3 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
| 4 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
|  | $\begin{aligned} & 100.000000 \\ & \text { year } \end{aligned}$ | 100.000000 | 100.000000 | 100.000000 | 100.000000 | 100.000000 |
| age | 2008 | 2009 |  |  |  |  |
| 1 | 2.127660 | 2.127660 |  |  |  |  |
| 2 | 3.571429 | 3.571429 |  |  |  |  |
| 3 | 100.000000 | 100.000000 |  |  |  |  |
|  | 100.00000 | 100.000 | 000 |  |  |  |
|  | 100.00000 | 100.000 | 000 |  |  |  |

Acoustic survey 1-9+ wr - Configuration
"Herring in Sub-area IV, Divisions VIId \& IIIa (autumn-spawners) . Imported from VPA file."
min max plusgroup minyear maxyear startf endf
$\begin{array}{lllll}1.00 & 9.00 & 9.00 & 1989.00 & 2008.00\end{array}$
Index type : number
Acoustic survey 1-9+ wr - Index Values
Units : NA

|  | ys : |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | 9361000 |
| 2 | 4090000 | 3306000 | 2634000 | 3734000 | 2984000 | 3185000 | 3849000 | 4497000 | 5960000 |
| 3 | 3903000 | 3521000 | 1700000 | 1378000 | 1637000 | 839000 | 2041000 | 2824000 | 2935000 |
| 4 | 1633000 | 3414000 | 1959000 | 1147000 | 902000 | 399000 | 672000 | 1087000 | 1441000 |
| 5 | 492000 | 1366000 | 1849000 | 1134000 | 741000 | 381000 | 299000 | 311000 | 601000 |
| 6 | 283000 | 392000 | 644000 | 1246000 | 777000 | 321000 | 203000 | 99000 | 215000 |
| 7 | 120000 | 210000 | 228000 | 395000 | 551000 | 326000 | 138000 | 83000 | 46000 |
| 8 | 44000 | 133000 | 94000 | 114000 | 180000 | 219000 | 119000 | 133000 | 78000 |
| 9 | 22000 | 43000 | 51000 | 104000 | 116000 | 131000 | 93000 | 206000 | 159000 |


| age | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 4449000 | 5087000 | 24736000 | 6837000 | 23055000 | 9829400 | 5183700 | 3114100 | 6822800 |
| 2 | 5747000 | 3078000 | 2923000 | 12290000 | 4875000 | 18949400 | 3415900 | 2055100 | 3772300 |
| 3 | 2520000 | 4725000 | 2156000 | 3083000 | 8220000 | 3081000 | 9191800 | 3648500 | 1997200 |
| 4 | 1625000 | 1116000 | 3140000 | 1462000 | 1390000 | 4188900 | 2167300 | 5789600 | 2097500 |
| 5 | 982000 | 506000 | 1007000 | 1676000 | 794600 | 675100 | 2590700 | 1212900 | 4175100 |
| 6 | 445000 | 314000 | 483000 | 450000 | 1031000 | 494800 | 317100 | 1174900 | 618200 |
| 7 | 170000 | 139000 | 266000 | 170000 | 244400 | 568300 | 327600 | 139900 | 562100 |
| 8 | 45000 | 54000 | 120000 | 98000 | 121000 | 145500 | 342050 | 126500 | 84300 |
| 9 | 121000 | 87000 | 97000 | 59000 | 149500 | 177700 | 185600 | 106700 | 70400 |

## year

age 20072008
162610003714000
227500002853000
318480001709000
48980001485000
5806000809000
61323000712000
$7 \quad 2430001749000$
$8 \quad 152000 \quad 185000$

Acoustic survey 1-9+ wr - Index Variance (Inverse Weights)

|  | ts : NA year |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 1 | 1.587302 | 1.587302 | 1.587302 | 1.587302 | 1.587302 | 1.587302 | 1.587302 |
| 2 | 1.612903 | 1.612903 | 1.612903 | 1.612903 | 1.612903 | 1.612903 | 1.612903 |
| 3 | 5.882353 | 5.882353 | 5.882353 | 5.882353 | 5.882353 | 5.882353 | 5.882353 |
| 4 | 10.000000 | 10.000000 | 10.000000 | 10.000000 | 10.000000 | 10.000000 | 10.000000 |
| 5 | 11.111111 | 11.111111 | 11.111111 | 11.111111 | 11.111111 | 11.111111 | 11.111111 |
| 6 | 12.500000 | 12.500000 | 12.500000 | 12.500000 | 12.500000 | 12.500000 | 12.500000 |
| 7 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 |
| 8 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 |
| 9 | 20.000000 | 20.000000 | 20.000000 | 20.000000 | 20.000000 | 20.000000 | 20.000000 |
|  | year |  |  |  |  |  |  |
| age | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 1 | 1.587302 | 1.587302 | 1.587302 | 1.587302 | 1.587302 | 1.587302 | 1.587302 |
| 2 | 1.612903 | 1.612903 | 1.612903 | 1.612903 | 1.612903 | 1.612903 | 1.612903 |
| 3 | 5.882353 | 5.882353 | 5.882353 | 5.882353 | 5.882353 | 5.882353 | 5.882353 |
| 4 | 10.000000 | 10.000000 | 10.000000 | 10.000000 | 10.000000 | 10.000000 | 10.000000 |
| 5 | 11.111111 | 11.111111 | 11.111111 | 11.111111 | 11.111111 | 11.111111 | 11.111111 |
| 6 | 12.500000 | 12.500000 | 12.500000 | 12.500000 | 12.500000 | 12.500000 | 12.500000 |
| 7 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 |
| 8 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 |
|  | $20.000000$ <br> year | 20.000000 | 20.000000 | 20.000000 | 20.000000 | 20.000000 | 20.000000 |
| age | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |  |
| 1 | 1.587302 | 1.587302 | 1.587302 | 1.587302 | 1.587302 | 1.587302 |  |
| 2 | 1.612903 | 1.612903 | 1.612903 | 1.612903 | 1.612903 | 1.612903 |  |
| 3 | 5.882353 | 5.882353 | 5.882353 | 5.882353 | 5.882353 | 5.882353 |  |
| 4 | 10.000000 | 10.000000 | 10.000000 | 10.000000 | 10.000000 | 10.000000 |  |
| 5 | 11.111111 | 11.111111 | 11.111111 | 11.111111 | 11.111111 | 11.111111 |  |
| 6 | 12.500000 | 12.500000 | 12.500000 | 12.500000 | 12.500000 | 12.500000 |  |
| 7 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 |  |
| 8 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 | 14.285714 |  |
| 9 | 20.000000 | 20.000000 | 20.000000 | 20.000000 | 20.000000 | 20.000000 |  |

Table 2.6.2.9 North Sea Autumn Spawning Herring. Stock object configuration

| min | max | plusgroup | minyear | maxyear | minfbar |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 9 | 9 | 1960 | 2008 | 2 |

Table 2.6.2.10 North Sea Autumn Spawning Herring. FLICA configuration settings

```
sep.2 : NA
sep.gradual : TRUE
sr : TRUE
sr.age : 1
lambda.age : 0.1 0.1 3.67 2.87 2.23 1.74 1.37 1.04 0.94 0
lambda.yr : 1 1 1 1 1
lambda.sr : 0.1
index.model : power linear linear linear
index.cor : F
sep.nyr : 5
sep.age : 4
sep.sel : 1
```

Table 2.6.2.11 North Sea Autumn Spawning Herring. FLR R software versions

R version 2.8.1 (2008-12-22)

```
Package : FLICA
Version : 1.4-9
Packaged : Sat Mar 21 13:07:02 2009; mpa
Built : R 2.8.0; ; 2009-03-21 13:07:03; windows
Package : FLAssess
Version : 1.99-102
Packaged : Wed Sep 24 16:58:05 2008; mpa
Built : R 2.7.0; i386-pc-mingw32; 2008-09-24 16:58:07; windows
Package : FLCore
Version : 3.0
Packaged : Tue Mar 10 04:42:26 2009; theussl
Built : R 2.8.1; i386-pc-mingw32; 2009-03-10 04:42:28; windows
```

Table 2.6.2.12 North Sea Autumn Spawning Herring. Stock Summary

| Year | Recruitment thousands Age 0 | TSB tonnes | SSB tonnes | Fbar $(2-6)$ | Fbar $(0-1)$ | Landings tonnes | SOP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 12087837 | 3723796.2 | 1861453 | 0.339 | 0.141 | 696200 | 1.18 |
| 1961 | 108858073 | 4343827.3 | 1643371 | 0.436 | 0.074 | 696700 | 1.13 |
| 1962 | 46275907 | 4385354.8 | 1103051 | 0.536 | 0.047 | 627800 | 1.17 |
| 1963 | 47657598 | 4611328.5 | 2172543 | 0.227 | 0.069 | 716000 | 0.86 |
| 1964 | 62784953 | 4783460.8 | 2018320 | 0.344 | 0.161 | 871200 | 1.07 |
| 1965 | 34894783 | 4332830.9 | 1438153 | 0.694 | 0.127 | 1168800 | 1.15 |
| 1966 | 27858148 | 3310425.9 | 1274731 | 0.620 | 0.103 | 895500 | 1.07 |
| 1967 | 40255855 | 2816512 | 921824 | 0.798 | 0.162 | 695500 | 1.18 |
| 1968 | 38698462 | 2520862.8 | 412347 | 1.336 | 0.168 | 717800 | 1.26 |
| 1969 | 21581503 | 1905375.6 | 424149 | 1.105 | 0.169 | 546700 | 0.97 |
| 1970 | 41072449 | 1921915.3 | 374690 | 1.106 | 0.152 | 563100 | 0.97 |
| 1971 | 32306362 | 1849387.7 | 266027 | 1.408 | 0.318 | 520100 | 1.07 |
| 1972 | 20858534 | 1549402.7 | 288267 | 0.697 | 0.318 | 497500 | 0.92 |
| 1973 | 10102036 | 1155879.9 | 233324 | 1.135 | 0.360 | 484000 | 0.96 |
| 1974 | 21688511 | 911745 | 161930 | 1.053 | 0.263 | 275100 | 0.97 |
| 1975 | 2814491 | 679921.7 | 81542 | 1.473 | 0.423 | 312800 | 0.93 |
| 1976 | 2720374 | 358115 | 77673 | 1.451 | 0.199 | 174800 | 0.95 |
| 1977 | 4326038 | 209911.7 | 47180 | 0.815 | 0.198 | 46000 | 1.20 |
| 1978 | 4594665 | 224320.4 | 64421 | 0.054 | 0.123 | 11000 | 1.22 |
| 1979 | 10600186 | 381521.6 | 106648 | 0.065 | 0.125 | 25100 | 1.01 |
| 1980 | 16716729 | 629858.8 | 130506 | 0.285 | 0.120 | 70764 | 1.09 |
| 1981 | 37860685 | 1158012.2 | 195088 | 0.353 | 0.384 | 174879 | 1.01 |
| 1982 | 64740217 | 1842378 | 277945 | 0.264 | 0.280 | 275079 | 0.98 |
| 1983 | 61794951 | 2717821.1 | 431973 | 0.338 | 0.326 | 387202 | 1.08 |
| 1984 | 53439842 | 2863511.4 | 678583 | 0.456 | 0.216 | 428631 | 1.05 |
| 1985 | 80893853 | 3460695.6 | 698559 | 0.644 | 0.234 | 613780 | 1.04 |
| 1986 | 97583821 | 3470896.9 | 678431 | 0.573 | 0.189 | 671488 | 1.14 |
| 1987 | 86180225 | 3934291.6 | 899264 | 0.553 | 0.267 | 792058 | 1.02 |
| 1988 | 42262187 | 3618637.4 | 1192606 | 0.539 | 0.353 | 887686 | 1.16 |
| 1989 | 39173461 | 3307320.2 | 1247533 | 0.547 | 0.281 | 787899 | 1.03 |
| 1990 | 35871987 | 2973565.7 | 1182522 | 0.443 | 0.256 | 645229 | 1.05 |
| 1991 | 33634923 | 2712272.5 | 978026 | 0.491 | 0.213 | 658008 | 1.02 |
| 1992 | 62138096 | 2433999 | 701221 | 0.584 | 0.342 | 716799 | 1.00 |
| 1993 | 50250988 | 2515705.5 | 470816 | 0.692 | 0.399 | 671397 | 1.02 |
| 1994 | 34500565 | 2022343.8 | 508453 | 0.709 | 0.236 | 568234 | 1.05 |
| 1995 | 41602863 | 1841589.4 | 460948 | 0.739 | 0.308 | 579371 | 1.01 |
| 1996 | 49747245 | 1623067.6 | 462153 | 0.402 | 0.165 | 275098 | 1.00 |
| 1997 | 28730419 | 1946578.6 | 560344 | 0.421 | 0.035 | 264313 | 1.00 |
| 1998 | 27373602 | 2066498.6 | 733708 | 0.486 | 0.088 | 391628 | 1.00 |
| 1999 | 67697220 | 2331905.9 | 857946 | 0.370 | 0.043 | 363163 | 1.00 |
| 2000 | 40678164 | 2858800.5 | 865448 | 0.362 | 0.062 | 388157 | 1.00 |
| 2001 | 90678376 | 3236723.2 | 1301315 | 0.294 | 0.051 | 374065 | 0.99 |
| 2002 | 30444152 | 3933139.9 | 1587054 | 0.244 | 0.040 | 394709 | 1.00 |
| 2003 | 19069558 | 3647277.7 | 1708980 | 0.245 | 0.063 | 482281 | 1.02 |
| 2004 | 23729100 | 3343287.6 | 1752878 | 0.298 | 0.062 | 587698 | 1.00 |
| 2005 | 16141706 | 2862232.2 | 1615686 | 0.366 | 0.075 | 663813 | 1.00 |
| 2006 | 27136239 | 2340941 | 1233800 | 0.349 | 0.072 | 514597 | 1.00 |
| 2007 | 17358063 | 2085121.8 | 952774 | 0.339 | 0.070 | 406482 | 1.01 |
| 2008 | 20044858 | 1868926.6 | 999336 | 0.236 | 0.049 | 257870 | 1.00 |

## Table 2.6.2.13 North Sea Autumn Spawning Herring. Estimated Fishing Mortality

| Units : year |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1960 | 1961 | 1962 | 1963 | 1964 | 965 |
| 0 | 0.02574284 | 0.01858824 | 0.004857388 | 0.01478961 | 0.01258623 | 0.007143512 |
| 1 | 0.25593593 | 0.12937978 | 0.089672658 | 0.12406122 | 0.30843530 | 0.246129893 |
| 2 | 0.43747076 | 0.61737576 | 0.250233880 | 0.29752416 | 0.38895838 | 0.775347360 |
| 3 | 0.32978964 | 0.35401289 | 0.627471177 | 0.27556155 | 0.41239120 | 0.738840142 |
| 4 | 0.34021201 | 0.41116964 | 0.424420432 | 0.22726359 | 0.37041932 | 0.776718126 |
| 5 | 0.26989927 | 0.40731634 | 0.539913509 | 0.15182178 | 0.30822633 | 0.660342541 |
| 6 | 0.31510911 | 0.38874213 | 0.836255146 | 0.18449853 | 0.24005044 | 0.520925638 |
| 7 | 0.62025784 | 0.25198971 | 0.656754412 | 0.29603392 | 0.28720584 | 0.462824553 |
| 8 | 0.59556157 | 0.54998866 | 0.586878207 | 0.35349764 | 0.59051441 | 0.890581695 |
|  | 0.59556157 | 0.54998866 | 0.586878207 | 0.35349764 | 0.59051441 | 0.890581695 |
| year |  |  |  |  |  |  |
| age | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 |
| 0 | 0.02145791 | 0.02563555 | 0.03481149 | 0.008238172 | 0.03510148 | 0.03397157 |
|  | 0.18523876 | 0.29805042 | 0.30024977 | 0.329105590 | 0.26806072 | 0.60218038 |
|  | 0.59208152 | 0.42223503 | 1.32730552 | 0.784407945 | 0.97284421 | 0.88261747 |
| 3 | 0.70824008 | 0.80462916 | 1.87222392 | 0.912609784 | 1.26704927 | 1.21478490 |
| 4 | 0.57175522 | 0.92443845 | 1.07164717 | 0.874419941 | 1.33130509 | 1.22679764 |
| 5 | 0.83467742 | 0.82752982 | 1.23399149 | 1.054647759 | 0.87625303 | 1.08703849 |
|  | 0.39078786 | 1.01011159 | 1.17445298 | 1.900945102 | 1.08154540 | 2.62891764 |
| 7 | 0.38954212 | 1.53682847 | 1.60168399 | 1.298798275 | 4.12205870 | 2.74329509 |
|  | 0.74863294 | 1.07648090 | 1.67681829 | 1.330363979 | 1.74277839 | 1.95979538 |
|  | 0.74863294 | 1.07648090 | 1.67681829 | 1.330363979 | 1.74277839 | 1.95979538 |
| year |  |  |  |  |  |  |


| age | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

00.058309110 .046199830 .074952120 .15826240 .14718470 .097789870 .04561793
10.578263540 .673970220 .452036030 .68843750 .25105080 .298329930 .20066532
20.812228161 .022376451 .028819361 .31426121 .34155280 .227382720 .02433139
30.801448131 .334182330 .973674931 .50536391 .44753331 .421767000 .04300972
40.799695180 .987946530 .994872981 .37632711 .74462500 .438870610 .10568299
50.549924570 .951671861 .186639331 .89309811 .61907291 .222475770 .01711883
60.520193461 .380914571 .079283511 .27794101 .1044608 0.76619614 0. 0.07941844
70.099771240 .813550520 .777439352 .03964171 .51701330 .797717170 .06447590
81.099596521 .619187851 .374311882 .08065221 .68336300 .989441310 .19933854
91.099596521 .619187851 .374311882 .08065221 .68336300 .989441310 .19933854 year

| age | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 0.08374700 | 0.12587185 | 0.4822165 | 0.3345558 | 0.3998236 | 0.2264703 | 0.08531661 |
| 1 | 0.16704921 | 0.11330386 | 0.2857669 | 0.2251992 | 0.2518700 | 0.2053172 | 0.38311217 |
| 2 | 0.09502297 | 0.36493691 | 0.3244910 | 0.2609733 | 0.3024553 | 0.3148049 | 0.40473190 |
| 3 | 0.06682723 | 0.42114976 | 0.2766902 | 0.5092484 | 0.3252089 | 0.4303415 | 0.67238411 |
| 4 | 0.09495012 | 0.29910506 | 0.3055837 | 0.2486825 | 0.4377708 | 0.5390206 | 0.73984869 |
| 5 | 0.05315146 | 0.26940314 | 0.4165956 | 0.1556973 | 0.2777677 | 0.6304763 | 0.66799319 |
| 6 | 0.01280347 | 0.06835163 | 0.4415147 | 0.1468505 | 0.3486088 | 0.3638644 | 0.73712230 |
| 7 | 0.45298473 | 0.10492787 | 0.9965562 | 0.2371677 | 0.3980662 | 0.7092367 | 0.56705127 |
| 8 | 0.24906938 | 0.38099618 | 0.6504193 | 0.4544292 | 0.5380664 | 0.6305582 | 0.89711388 |
| 9 | 0.24906938 | 0.38099618 | 0.6504193 | 0.4544292 | 0.5380664 | 0.6305582 | 0.89711388 | year

$\begin{array}{llllllll}\text { age } & 1986 & 1987 & 1988 & 1989 & 1990 & 1991 & 1992\end{array}$
00.061957760 .16144040 .12473020 .13031170 .058877830 .11785470 .2967334
10.315982830 .37248590 .58034730 .43108870 .452705730 .30820050 .3872125
20.460010280 .40668220 .35592030 .39872060 .377350220 .57415850 .5727086
30.523433150 .50662300 .40146640 .41053770 .370090240 .45539570 .4981599
40.583908290 .59091600 .58415810 .55736580 .468243290 .45873210 .5743451
50.557115850 .62036840 .66804870 .66117780 .502478220 .48473350 .5484874
60.741404860 .64228580 .68291160 .70882600 .498682680 .48175890 .7248097
70.834505590 .62712370 .70559050 .72439050 .696026770 .43130760 .7068996
80.835953980 .82620870 .96237060 .86884890 .808621500 .74001620 .8970386
90.835953980 .82620870 .96237060 .86884890 .808621500 .74001620 .8970386 year
$\begin{array}{llllllll}\text { age } & 1993 & 1994 & 1995 & 1996 & 1997 & 1998 & 1999\end{array}$
00.37610260 .22630230 .32233630 .075848060 .023980340 .015093340 .03717430
10.42213880 .24604240 .29314710 .254253300 .045590940 .160099510 .04941211
20.66855870 .68370890 .59989890 .308621920 .287592500 .268238780 .24745134
30.64084130 .71623500 .86695050 .490244930 .391823160 .408561390 .42068092
40.73288610 .91166120 .86692260 .418312950 .511724000 .424724950 .38869502
50.71513990 .55669190 .82406930 .478003540 .452636800 .607544730 .33867430
60.70434840 .67658540 .53857460 .314700760 .460826660 .721911780 .45348849
70.88672840 .48242570 .65819330 .142665900 .241755380 .377109360 .46052795
81.04654570 .88501180 .95727870 .558163350 .419438470 .557166620 .41254439
91.04654570 .88501180 .95727870 .558163350 .419438470 .557166620 .41254439

| year |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 0 | 0.04376704 | 0.03240791 | 0.03856206 | 0.0310161 | 0.05907963 | 0.07243465 | 0.06920000 |
| 1 | 0.07984775 | 0.06986187 | 0.04174734 | 0.0941856 | 0.06407668 | 0.07856128 | 0.07505304 |
| 2 | 0.23115453 | 0.12703215 | 0.14621638 | 0.1320715 | 0.14985609 | 0.18373123 | 0.17552652 |
| 3 | 0.31705738 | 0.30256794 | 0.22394358 | 0.2046686 | 0.23445697 | 0.28745623 | 0.27461957 |
| 4 | 0.46656992 | 0.30386150 | 0.28498251 | 0.2899446 | 0.32991277 | 0.40448992 | 0.38642701 |
| 5 | 0.47937736 | 0.39604002 | 0.22079729 | 0.3639487 | 0.38348993 | 0.47017826 | 0.44918196 |
| 6 | 0.31597906 | 0.34227126 | 0.34461327 | 0.2345661 | 0.39294747 | 0.48177369 | 0.46025959 |
| 7 | 0.39585030 | 0.21149488 | 0.27809321 | 0.3871822 | 0.34856900 | 0.42736342 | 0.40827906 |
| 8 | 0.41215408 | 0.31389006 | 0.26849109 | 0.3185907 | 0.32991277 | 0.40448992 | 0.38642701 |
| 9 | 0.41215408 | 0.31389006 | 0.26849109 | 0.3185907 | 0.32991277 | 0.40448992 | 0.38642701 |
| year |  |  |  |  |  |  |  |
| age | 2007 | 2008 |  |  |  |  |  |
| 0 | 0.06712092 | 0.04679468 |  |  |  |  |  |
| 1 | 0.07279811 | 0.05075265 |  |  |  |  |  |
| 2 | 0.17025290 | 0.11869519 |  |  |  |  |  |
| 3 | 0.26636875 | 0.18570426 |  |  |  |  |  |
| 4 | 0.37481698 | 0.26131109 |  |  |  |  |  |
| 5 | 0.43568649 | 0.30374747 |  |  |  |  |  |
| 6 | 0.44643129 | 0.31123843 |  |  |  |  |  |
| 7 | 0.39601250 | 0.27608797 |  |  |  |  |  |
| 8 | 0.37481698 | 0.26131109 |  |  |  |  |  |
| 9 | 0.37481698 | 0.26131109 |  |  |  |  |  |

Table 2.6.2.14 North Sea Autumn Spawning Herring. Estimated Population Abundance


| year |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1994 | 1995 | 1996 | 1997 | 1998 | 81999 |
| 0 | 34500564.63 | 41602862.72 | 49747244.618 | 28730419.08 | 27373601.838 | 867697219.697 |
| 1 | 12691438.40 | 10121616.83 | 11087659.044 | 16964230.16 | 10318889.224 | 49919333.955 |
| 2 | 4097951.17 | 3650577.11 | 2777430.711 | 3163184.87 | 5962656.369 | 9 3234507.286 |
| 3 | 1042830.84 | 1532312.87 | 1484363.940 | 1511200.51 | 1757666.177 | 73377977.985 |
| 4 | 606123.45 | 417155.56 | 527201.639 | 744339.10 | 836173.805 | 5956404.349 |
| 5 | 267075.87 | 220395.25 | 158623.998 | 313961.43 | 403740.648 | 8494779.169 |
| 6 | 195116.77 | 138495.92 | 87475.019 | 88990.81 | 180662.988 | 198984.709 |
| 7 | 208936.33 | 89748.64 | 73132.015 | 57780.57 | 50790.352 | 279417.724 |
| 8 | 124343.80 | 116699.69 | 42048.324 | 57374.52 | 41054.363 | 3 31519.206 |
| 9 | 82777.45 | 49494.35 | 9654.364 | 17414.10 | 9265.759 | 96363.323 |
| year |  |  |  |  |  |  |
| age | 2000 | 02001 | 2002 | 2003 | 2004 | 2005 |
| 0 | 40678163.506 | 90678376.18 | 30444151.68 | 19069557.74 | 23729099.51 | 16141705.8 |
| 1 | 23995607.883 | 14323827.15 | 32294954.42 | 10776112.25 | 6801050.8 | 8228654.3 |
| 2 | 3473190.639 | 8150041.84 | 4913873.43 | 11394874.69 | 3607973.5 | 2346677.4 |
| 3 | 1870910.611 | 12041978.39 | 5317434.27 | 3145101.41 | 7397130.9 | 2300876.6 |
| 4 | 1815927.538 | 1115572.68 | 1235346.13 | 3480052.33 | 2098405.0 | 4790502.1 |
| 5 | 586683.662 | 21030481.80 | 744908.71 | 840606.76 | 2356323.9 | 1365150.6 |
| 6 | 319078.529 | 9328687.82 | 627498.77 | 540483.65 | 528570.0 | 1452974.5 |
| 7 | 114404.395 | 5210494.22 | 211206.67 | 402271.99 | 386796.3 | 322862.9 |
| 8 | 45340. 180 | 9 69678.32 | 154155.73 | 144711.76 | 247137.8 | 246985.4 |
| 9 | 7636.273 | $3 \quad 10473.15$ | 38748.21 | 34742.19 | 118592.3 | 184153.0 |
| year ${ }^{\text {a }}$ |  |  |  |  |  |  |
| age | 2006 | 2007 | 2008 |  |  |  |
| 0 | 27136239.01 | 17358062.7020 | 20044857.8 |  |  |  |
| 1 | 5523278.9 | 9315410.43 | 5971130.0 |  |  |  |
| 2 | 2798437.5 | 1884982.72 | 3186336.9 |  |  |  |
| 3 | 1446677.0 | 1739390.38 | 1177820.6 |  |  |  |
| 4 | 1413167.9 | 900009.15 | 1091077.7 |  |  |  |
| 5 | 2892569.8 | 868842.79 | 559804.7 |  |  |  |
| 6 | 771889.8 | 1670233.35 | 508505.3 |  |  |  |
| 7 | 812076.4 | 440796.00 | 967085.9 |  |  |  |
| 8 | 190540.4 | 488488.21 | 268424.5 |  |  |  |
| 9 | 141757.2 | 78293.18 | 195699.5 |  |  |  |

Table 2.6.2.15 North Sea Autumn Spawning Herring. Survivors after terminal year.

| Units <br> year | $N A$ |
| :--- | ---: |
| age | 2009 |
| 0 | 32832169.0 |
| 1 | 7036972.0 |
| 2 | 2087951.7 |
| 3 | 2096306.0 |
| 4 | 800884.5 |
| 5 | 760221.7 |
| 6 | 373844.7 |
| 7 | 337052.0 |
| 8 | 663945.0 |
| 9 | 323384.1 |

Table 2.6.2.16 North Sea Autumn Spawning Herring. Fitted Selection Pattern

| Units <br> year <br> age | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 0 | 0.1790765 | 0.1790765 | 0.1790765 | 0.1790765 | 0.1790765 |
| 1 | 0.1942231 | 0.1942231 | 0.1942231 | 0.1942231 | 0.1942231 |
| 2 | 0.4542294 | 0.4542294 | 0.4542294 | 0.4542294 | 0.4542294 |
| 3 | 0.7106635 | 0.7106635 | 0.7106635 | 0.7106635 | 0.7106635 |
| 4 | 1.0000000 | 1.0000000 | 1.0000000 | 1.0000000 | 1.0000000 |
| 5 | 1.1623980 | 1.1623980 | 1.1623980 | 1.1623980 | 1.1623980 |
| 6 | 1.1910647 | 1.1910647 | 1.1910647 | 1.1910647 | 1.1910647 |
| 7 | 1.0565490 | 1.0565490 | 1.0565490 | 1.0565490 | 1.0565490 |
| 8 | 1.0000000 | 1.0000000 | 1.0000000 | 1.0000000 | 1.0000000 |
| 9 | 1.0000000 | 1.0000000 | 1.0000000 | 1.0000000 | 1.0000000 |

Table 2.6.2.17 North Sea Autumn Spawning Herring. Predicted catch in numbers

| Unitsyear |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1960 | 1961 | 1962 | 2 1963 | 1964 | 1965 | 1966 | 1967 |  | 968 |
| 0 | 194600 | 1269200 | 141800 | 442800 | 496900 | 157100 | 374500 | 645400 | 8393 | 300 |
| 1 | 2392700 | 336000 | 2146900 | 1262200 | 2971700 | 32093001 | 1383100 | 1674300 | 24250 | 000 |
| 2 | 1142300 | 1889400 | 269600 | 2961200 | 1547500 | 22176002 | 2569700 | 1171500 | 17952 | 00 |
| 3 | 1966700 | 479900 | 797400 | 177200 | 2243100 | 1324600 | 741200 | 1364700 | 14943 |  |
| 4 | 165900 | 1455900 | 335100 | 158300 | 148400 | 2039400 | 450100 | 371500 | 621 | 400 |
| 5 | 167700 | 124000 | 1081800 | 80600 | 149000 | 145100 | 889800 | 297800 | 1571 |  |
| 6 | 112900 | 157900 | 126900 | 229700 | 95000 | 151900 | 45300 | 393100 | 1450 | 00 |
| 7 | 125800 | - 61400 | 145100 | - 22400 | 256300 | 117600 | 64800 | 67900 | 163 | 400 |
| 8 | 128600 | - 56000 | 86300 | 42000 | 26300 | 413000 | 95500 | 81600 |  | 00 |
| 9 | 142000 | - 87500 | - 86800 | - 51000 | 57700 | 78400 | 236300 | 172800 | 918 |  |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 |
| 0 | 112000 | 898100 | 684000 | 750400 | 289400 | 9961002 | 26380023 | 238200256 | 56800 | 130000 |
| 1 | 2503300 | 1196200 | 4378500 | 3340600 | 2368000 | 84610024 | 246050012 | 126600144 | 44300 | 168600 |
| 2 | 1883000 | 2002800 | 1146800 | 1440500 | 1344200 | 772600 | 54170090 | 90150044 | 44700 | 4900 |
| 3 | 296300 | 883600 | 662500 | 343800 | 659200 | 362000 | 25960011 | 117300186 | 86400 | 5700 |
| 4 | 133100 | 125200 | 208300 | 130600 | 150200 | 1260001 | 140500 | 5200010 | 10800 | 5000 |
| 5 | 190800 | - 50300 | - 26900 | 32900 | 59300 | 56100 | 57200 | 34500 | 7000 | 300 |
| 6 | 49900 | - 61000 | 30500 | - 5000 | 30600 | 22300 | 16100 | 6100 | 4100 | 200 |
| 7 | 42700 | 7900 | - 26800 | - 200 | 3700 | 5000 | 9100 | 4400 | 1500 | 200 |
| 8 | 27400 | 12000 | 100 | 1100 | 1400 | 2000 | 3400 | 1000 | 700 | 200 |
| 9 | 25100 | - 12200 | 12400 | - 400 | 600 | 1100 | 1400 | 400 | - | 300 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |  | 1987 |
| 0 | 542000 | 1262700 | 9519700 | 11956700 | 13296900 | 6973300 | 4211000 | 3724700 | 08229 | 200 |
| 1 | 159200 | 245100 | 872000 | 1116400 | 2448600 | 1818400 | 3253000 | 4801400 | 06836 | 3300 |
| 2 | 34100 | 134000 | 284300 | 299400 | 573800 | 1146200 | 1326300 | 1266700 | 02137 | 200 |
| 3 | 10000 | 91800 | 56900 | 230100 | 216400 | 441400 | 1182400 | - 840800 |  | 900 |
| 4 | 10100 | 32200 | 39500 | 33700 | 105100 | 201500 | - 368500 | - 465900 |  | 100 |
| 5 | 2100 | 21700 | 28500 | 14400 | 26200 | - 81100 | 124500 | 129800 |  | 800 |
| 6 | 200 | 2300 | 22700 | 6800 | 22800 | 22600 | - 43600 | - 62100 |  | 4700 |
| 7 | 800 | 1400 | 18700 | 7800 | 12800 | 25200 | - 20200 | - 20500 |  | 800 |
| 8 | 600 | 400 | 5500 | 3600 | 11000 | 11100 | - 13100 | - 13600 |  | 8000 |
| 9 | 100 | 100 | 1100 | 1100 | 12100 | 18600 | - 16100 | - 14800 |  | 8200 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 95 | 1996 |
| 0 | 3164800 | 3057800 | 1302800 | 2386600 | 10331300 | 10265400 | 4498900 | 00 7438469 |  | 11226 |
| 1 | 7867000 | 3145900 | 3020000 | 2138900 | 2303100 | 3826800 | 1785200 | 1664874 | 74160 | 06393 |
| 2 | 2232500 | 1593700 | - 899300 | 1132800 | 1284900 | 1176300 | 1783200 | 001444061 |  | 42084 |
| 3 | 1090700 | 1363800 | 779100 | 556700 | 442700 | - 609000 | 489100 | 00 816703 |  | 25601 |
| 4 | 383700 | 809300 | 861000 | 548900 | 361500 | 305500 | 347600 | 00231794 |  | 12099 |
| 5 | 255800 | 211800 | 387500 | 501200 | 360500 | - 215600 | 109000 | 00118536 |  | 57586 |
| 6 | 128100 | 123700 | - 80200 | 205300 | 375600 | - 226000 | 91800 | 55128 |  | 22534 |
| 7 | 38000 | 61000 | 54400 | 39300 | 152400 | 188000 | 76400 | 41409 |  | 9264 |
| 8 | 15300 | 19500 | 28800 | 25600 | 39200 | - 87300 | 70000 | 68955 |  | 17195 |
| 9 | 8500 | 8700 | 11900 | 13000 | 23300 | 41700 | 46000 | 029245 |  | 3948 |


| year |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| age | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| 0 | 431175 | 259526 | 1566349 | 1105085 | 1832691 | 730279 | 369074 | 864675.83 | 717192.41 |
| 1 | 479702 | 977680 | 303520 | 1171677 | 614469 | 837557 | 617021 | 268233.90 | 395531.34 |
| 2 | 687920 | 1220105 | 616354 | 622853 | 842635 | 579592 | 1221992 | 435420.21 | 341838.94 |
| 3 | 446909 | 537932 | 1058716 | 463170 | 485628 | 970577 | 529386 | 1406680.78 | 523487.78 |
| 4 | 284920 | 276333 | 294066 | 646814 | 278884 | 292205 | 835552 | 562695.51 | 1521724.16 |
| 5 | 109178 | 175817 | 135648 | 213466 | 321743 | 140701 | 244780 | 716510.09 | 489212.51 |
| 6 | 31389 | 88927 | 69299 | 82481 | 90918 | 174570 | 107751 | 163977.08 | 530735.30 |
| 7 | 11832 | 15232 | 27998 | 35706 | 38252 | 48908 | 123291 | 108642.71 | 107231.16 |
| 8 | 18770 | 16766 | 10174 | 14624 | 17910 | 34620 | 37671 | 66270.98 | 78456.01 |
| 9 | 5697 | 3784 | 2054 | 2463 | 2692 | 8702 | 9044 | 31801.00 | 58497.00 |
| year |  |  |  |  |  |  |  |  |  |
| age | 2006 | 2007 | 2008 |  |  |  |  |  |  |
| 0 | 1153386.99 | 716227.5 | 581489.23 |  |  |  |  |  |  |
| 1 | 254001.62 | 415907.7 | 187562.29 |  |  |  |  |  |  |
| 2 | 390915.84 | 256024.8 | 309011.00 |  |  |  |  |  |  |
| 3 | 316307.44 | 370286.9 | 181482.64 |  |  |  |  |  |  |
| 4 | 432422.39 | 268554.9 | 239285.05 |  |  |  |  |  |  |
| 5 | 999762.42 | 293071.0 | 139901.50 |  |  |  |  |  |  |
| 6 | 271997.29 | 574467.5 | 129761.37 |  |  |  |  |  |  |
| 7 | 259924.68 | 137620.1 | 222537.16 |  |  |  |  |  |  |
| 8 | 58304.43 | 145760.6 | 58868.37 |  |  |  |  |  |  |
| 9 | 43377.00 | 23362.0 | 42919.00 |  |  |  |  |  |  |

Table 2.6.2.18 North Sea Autumn Spawning Herring. Catch Residuals

| UnitsU <br> year <br> age thousands NA | 2004 | 2005 | 2006 |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | -0.18923751 | 0.34784540 | -0.272086254 | -0.1426587551748559057341 | 0.31687197 |
| 1 | -0.26084251 | 0.59281726 | -0.134163373 | -0.5685273456876875908250 | 0.22556813 |
| 2 | 0.02829862 | 0.03905334 | 0.025686068 | -0.1556776339151793864257 | 0.07107111 |
| 3 | -0.02923265 | -0.07482783 | -0.018202293 | 0.1198915057863032262997 | 0.01795722 |
| 4 | -0.03493712 | -0.14323780 | 0.071817008 | 0.0620480828058247377288 | -0.18400401 |
| 5 | 0.04997962 | -0.01909213 | -0.001982850 | 0.0543946545494838409884 | -0.01710378 |
| 6 | 0.03208736 | 0.08211159 | -0.075767948 | 0.0909849863773814493406 | -0.09205924 |
| 7 | -0.03462823 | 0.07178706 | -0.050833542 | 0.0715662848336484219702 | -0.03193832 |
| 8 | -0.01413238 | 0.11832661 | 0.078012075 | -0.0887034451925809569461 | 0.23333173 |
| 9 | 0.00000000 | 0.00000000 | 0.000000000 | -0.0000000000000001110223 | 0.00000000 |

Table 2.6.2.19 North Sea Autumn Spawning Herring. Predicted Index Values
MLAI
Units : NA NA

| age | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | all 17.1829511 .235415 .0586944 .7805772 .6770223 .8458416 .9123598 .741913 year


| age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

all $13.9536821 .0617535 .17468 \quad 59.4804561 .5218659 .4649182 .529114 .6086$ year

| age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | all 120.7709113 .482390 .9950961 .7946438 .8799642 .5180937 .9337238 .04904 year

$\begin{array}{lllllllll}\text { age } & 1997 & 1998 & 1999 & 2000 & 2001 & 2002 & 2003 & 2004\end{array}$ all 47.60665 .1370478 .1351778 .93044126 .8478159 .7921174 .1584179 .3726 year
age $2005 \quad 2006 \quad 2007 \quad 2008$ all 163.1501119 .225988 .268293 .30518

MIK 0-wr
Units : NA NA

| age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | 0175.7339140 .712898 .43448117 .2819144 .629884 .0710480 .18975197 .7694 year

$\begin{array}{lllllllll}\text { age } 2000 & 2001 & 2002 & 2003 & 2004 & 2005 & 2006 & 2007\end{array}$
0118.7385265 .064088 .9235455 .7523469 .1321846 .9486978 .9585750 .52005 year
age 20082009
058.4882995 .8

IBTS1: 1-5+ wr
Units : NA NA year
$\begin{array}{lllllll}\text { age } & 1984 & 1985 & 1986 & 1987 & 1988 & 1989\end{array}$ 12278.055882291 .453424028 .170774939 .095433847 .585821994 .24484 $2 \quad 666.34730 \quad 617.26233 \quad 527.77901 \quad 990.498061163 .26128 \quad 751.45978$ $\begin{array}{llllll}128.16866 & 237.03349 & 206.77210 & 168.80319 & 336.32244 & 412.45535\end{array}$ $\begin{array}{lllllll}4 & 29.22657 & 41.47482 & 63.28486 & 62.83536 & 52.10119 & 114.19252 \\ 5 & 11.88516 & 14.13035 & 16.63513 & 25.19389 & 29.17576 & 27.55435\end{array}$ year

| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

$\begin{array}{llllllll}1 & 1833.24675 & 1835.91461 & 1606.88517 & 2471.55511 & 1887.32658 & 1496.33535\end{array}$ $\begin{array}{llllll}445.01582 & 391.66898 & 445.18705 & 359.29100 & 534.97944 & 481.59460\end{array}$ $\begin{array}{llllll}257.95072 & 153.99138 & 113.48277 & 126.86989 & 93.27080 & 134.49230\end{array}$ $\begin{array}{rrrrrrr}4 & 140.49141 & 91.13732 & 49.76946 & 34.63690 & 33.42277 & 23.13173 \\ 5 & 45.37544 & 67.68483 & 64.93911 & 44.45886 & 27.48521 & 18.93528\end{array}$ year
$\begin{array}{lllllll}\text { age } & 1996 & 1997 & 1998 & 1999 & 2000 & 2001\end{array}$ 11647.139292586 .736761551 .081881511 .795863643 .259812177 .50700 $2 \quad 379.99319 \quad 433.90912 \quad 819.90708 \quad 445.92464 \quad 479.807091140 .64460$ $\begin{array}{lrrrrrr}3 & 136.56534 & 140.75545 & 163.36945 & 313.49696 & 175.89559 & 192.32672\end{array}$
$\begin{array}{rrrrrrr}4 & 30.92007 & 43.14829 & 49.00182 & 56.30061 & 105.86259 & 66.37045 \\ 5 & 11.98965 & 17.21664 & 21.52126 & 26.21995 & 34.53573 & 53.49843\end{array}$ year
$\begin{array}{llllllll}\text { age } 2002 & 2003 & 2004 & 2005 & 2006 & 2007 & 2008\end{array}$ 14926.760051633 .210401034 .64311249 .5598839 .103731415 .60824909 .90153 $2 \quad 686.077261593 .77312 \quad 503.5178 \quad 326.1113 \quad 389.29024 \quad 262.39259446 .41091$ $3 \quad 505.77678 \quad 299.87329 \quad 702.6666 \quad 217.1212136 .73425 \quad 164.57001112 .56718$ $\begin{array}{lrrrrrrrr}4 & 73.66996 & 207.40450 & 124.4377 & 281.4460 & 83.21248 & 53.07279 & 65.25932\end{array}$ $\begin{array}{llllllll}58.19482 & 63.87802 & 117.6825 & 114.3155 & 154.34294 & 114.02669 & 81.77310\end{array}$ year
age 2009
11072.31824
292.52538
200. 34907
47.90234
$5 \quad 80.58949$

Acoustic survey 1-9+ wr

| NA NA |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 1 | NA | NA | NA | NA | NA | NA | NA |
| 2 | 5800435.14 | 3466368.51 | 2806031.46 | 3191416.64 | 2472838.6 | 3658391.5 | 3412745.8 |
| 3 | 5624744.54 | 3578723.94 | 2060357.55 | 1491018.10 | 1568830.7 | 1116985.2 | 1510709.0 |
| 4 | 2468798.89 | 3154624.05 | 2054705.70 | 1068260.09 | 695008.9 | 621579.1 | 438449.6 |
| 5 | 555220.24 | 1359389.26 | 1825835.76 | 1153172.15 | 519302.1 | 343238.9 | 244510.7 |
| 6 | 293928.57 | 277097.56 | 735476.41 | 871162.01 | 540703.5 | 229375.7 | 175652.9 |
| 7 | 131397.63 | 122360.73 | 146667.25 | 337091.80 | 324208.3 | 253662.2 | 98920.3 |
| 8 | 35311.91 | 56468.02 | 55307.95 | 68502.54 | 128100.4 | 124184.2 | 112008.3 |
| 9 | 38263.82 | 56668.22 | 68213.98 | 98891. 51 | 148612.3 | 200787.5 | 115376.8 |
| year |  |  |  |  |  |  |  |
| age | 1996 | 1997 | 1998 | 1999 |  | 000 | 2001 |
| 1 | NA | 11219244.20 | 6407825.14 | 6546350.03 | 15573224 | . 799347406 | 6.17 |
| 2 | 3047618.74 | 3511277.12 | 6689646.67 | 3670598.32 | 23976949 | . 159882165 | 5.64 |
| 3 | 1800339.44 | 1934841.12 | 2229777.29 | 4256836.90 | 02495944 | 692745958 | 8.92 |
| 4 | 709176.28 | 951121.36 | 1120836.91 | 1307656.24 | 42378752 | . 241598134 | 4.84 |
| 5 | 212875.87 | 427260.75 | 504565.51 | 716884.86 | 6786743 | . 911446691 | 1.20 |
| 6 | 125480.79 | 117797.05 | 207154.87 | 264461. 09 | 9457388 | . 70464398 | 8. 97 |
| 7 | 107029.87 | 80077.50 | 65339.99 | 97586.50 | 0145667 | . 96296617 | 7.65 |
| 8 | 50264.63 | 74023.39 | 49103.36 | 40819.90 | 058731 | . 7795270 | 0.64 |
| 9 | 28029.80 | 54567.42 | 26916.30 | 20015.35 | 524024 | . 4734779 | 9.32 |
| year |  |  |  |  |  |  |  |
| age | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| 1 | 21403372.5 | 6938796.2 | 4452356.4 | 5344202.5 | 3594090.8 | 6069217.2 | 3937794.4 |
| 2 | 5895679.6 | 13778379.6 | 4320200.0 | 2758051.9 | 3303881.3 | 2231905.9 | 3881283.5 |
| 3 | 7466643.1 | 4463360.6 | 10327027.1 | 3119937.01 | 1975560.3 | 2386088.6 | 1689025.9 |
| 4 | 1788190.1 | 5023721.3 | 2963344.9 | 6493222.5 | 1934583.8 | 1239977.8 | 1600055.0 |
| 5 | 1151588.7 | 1201140.5 | 3330951.0 | 1839955.3 | 3943898.9 | 1193456.7 | 826831.9 |
| 6 | 885444.0 | 810246.0 | 726282.5 | 1901270.0 | 1022068.5 | 2228460.4 | 730828.9 |
| 7 | 286917.2 | 514649.6 | 505472.3 | 404028.71 | 1026950.3 | 561203.3 | 1315203.8 |
| 8 | 216105.2 | 197352.4 | 334945.0 | 321286.0 | 250335.4 | 645895.7 | 377783.2 |
| 9 | 131928.8 | 115074.3 | 390367.2 | 581810.7 | 452337.6 | 251428.6 | 668948.6 |

Table 2.6.2.20 North Sea Autumn Spawning Herring. Index Residuals

MLAI


| IBTS1: 1-5+ wr |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units : NA year |  |  |  |  |  |  |
| age | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 1 | -0.4074932 | -0.08854503 | -0.4139297 - | -0.29075254 | 0.1328323 | 0.1562747 |
| 2 | -1.4174297 | 0.15624061 | 0.3933331 - | -0.07650083 | 1.2751007 | 0.1525901 |
| 3 | -0.7354813 | 0.17367018 | 0.2888959 - | -0.37243145 | 0.8558949 | -0.1975625 |
| 4 | -0.0833981 | 0.01467620 | 0.2218904 - | -0.05704904 | 0.1065135 | -0.2449777 |
| 5 | -0.1491846 | 0.68176983 | 0.5233982 | 0.66037034 | -0.1523048 | -1.1738615 |
| year |  |  |  |  |  |  |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| 1 | -0.54633775 | -0.3554255 | -0.23674243 | $0.1226292-$ | -0.07437308 | -0.1060646 |
| 2 | 0.03766500 | 0.5706418 | -0.01956183 | 0.7846304 | 0.78015268 | 1.0620096 |
| 3 | 0.08123504 | 0.5184013 | 0.53147947 | 0.5621470 | 0.82606205 | 0.7280691 |
| 4 | 0.64996647 | 0.8881591 | 0.09647116 | 0.2626758 | 0.72492697 | 0.4607536 |
| 5 | 0.45196734 | 0.7694019 | 0.35124735 | 0.3880048 | 0.43592655 | -1.0614092 |
| year |  |  |  |  |  |  |
| age | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | 0.1379981 | 0.5322630 | 0.38339874 - | -0.69717159 | 0.02222319 | 0.13786646 |
| 2 | -0.6543034 | 0.1547603 - | -0.03514365 | 0.01048110 | -0.87820167 | -0.00998849 |
| 3 | -1.4240907 | 0.1446382 - | -0.53521396 | 0.46946493 | -0.12846099 | 0.49992380 |
| 4 | -1.1097042 | -0.3458674 - | -0.85642371 | 0.55608696 | -0.58735398 | 0.34682933 |
| 5 | -0.3947759 | 0.1466005 - | -0.18754221 | 0.30486937 | -1.35011738 | 0.24402316 |
| year |  |  |  |  |  |  |
| age | 2002 | 2003 | 2004 | 42005 | 2006 | 2007 |
| 1 | -0.19230991 | 0.52651433 | -0.05517701 | -0.2212076 | 6 0.08247307 | -0.06916648 |
| 2 | -0.04155130 | -0.02393311 | -0.14278712 | $2-0.4203707$ | 1.33637043 | -1.65775882 |
| 3 | -0.40259698 | 0.71319249 | 0.08634008 | 80.4944948 | 10.89529096 | -2.19917099 |
| 4 | -1.07880115 | 0.55140247 | -0.10197259 | 90.3234381 | 1.06396793 | -1.90012891 |
| 5 | -1.07126914 | -0.18355624 | 40.37366147 | 70.1398998 | 80.75509799 | -1.01595751 |
| year |  |  |  |  |  |  |
| age | 2008 | 2009 |  |  |  |  |
| 1 | 0.7369741 | 0.7832426 |  |  |  |  |
| 2 | -0.5692585 | -0.7671915 |  |  |  |  |
| 3 | 0.3211000 | -2.1952952 |  |  |  |  |
| 4 | 1.3623181 | -1.2644027 |  |  |  |  |
| 5 | 0.8957181 | -0.3819805 |  |  |  |  |

Acoustic survey 1-9+ wr

|  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1989 | 1990 | 1991 | 1992 | 21993 | 31994 |
| 1 | NA | NA | A NA | A NA | A NA | NA NA |
| 2 | -0.34938797 | -0.047368507 | -0.06326759 | 90.15701514 | 40.18789798 | -0.1385713 |
| 3 | -0.36543004 | -0.016261254 | -0.19225129 | -0.07882601 | 10.04253472 | - 0.2861779 |
| 4 | -0.41331294 | 0.079015294 | -0.04769849 | 90.07111860 | 0.26068982 | -0.4433018 |
| 5 | -0.12088614 | 0.004851235 | 50.01260712 | $2-0.01676533$ | 30.35551492 | 20.1043726 |
| 6 | -0.03788990 | 0.346892189 | -0.13281974 | $4 \quad 0.35786574$ | $4 \quad 0.36256920$ | 20.3360801 |
| 7 | -0.09073632 | 0.540134030 | 0.0 .44117924 | 40.15853045 | $5 \quad 0.53034872$ | 20.2508940 |
| 8 | 0.21996933 | 0.856674606 | 60.53037820 | 0.50932766 | 60.34014248 | 480.5673059 |
|  | $-0.55346243$ <br> year | -0.276013431 | - -0.29082396 | 60.05036748 | $8-0.24775086$ | -0.4270498 |
| age | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 1 | NA | NA | -0.18107841 - | -0.36484057 | -0.2522194 0 | 0.46270659 |
| 2 | 0.12029620 | 0.3890500 | 0.52909066 | -0.15188308 | -0.1760746-0 | 0.30790449 |
| 3 | 0.30086078 | 0.4501791 | 0.41668224 | 0.12235719 | $0.1043412-0$ | 0.14641264 |
| 4 | 0.42701340 | 0.4270728 | 0.41545093 | 0.37143217 | -0.1584855 0 | 0.27764672 |
| 5 | 0.20118459 | 0.3790837 | 0.34120046 | 0.66589363 | -0.3483786 0 | 0.24682810 |
| 6 | 0.14469633 | -0.2370328 | 0.60167480 | 0.76460759 | 0.17169890 | 0.05448308 |
| 7 | 0.33293923 | -0.2542673 | -0.55435358 | 0.95619422 | 0.35373480 | 0.60216654 |
| 8 | 0.06055088 | 0.9730475 | 0.05232767 | -0.08726495 | 0.27981440 | 0.71451090 |
|  | -0.21560363 | 1.9946080 | 1.06946717 | 1.50305847 | 1.46940881 | 1.39563802 |
|  | year |  |  |  |  |  |
| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | -0.31274985 | 0.07433416 | 0.34824959 | 0.15208560 | -0.5400721 | 0.64097889 |
| 2 | 0.21805424 | -0.19009971 | 0.31867160 | -0.23486070 | -0.2942001 | 0.13258696 |
| 3 | 0.11577280 | 0.09612469 | -0.37064776 | -0.11645267 | 0.1565033 | 0.01089416 |
| 4 | -0.08903186 | -0.25190023 | -0.18173279 | -0.31283651 | -0.1146957 | 0.08085392 |
| 5 | 0.14713099 | -0.37105888 | -0.57616594 | -0.25132973 | -0.4167271 | 0.05696851 |
| 6 | -0.03149645 | 0.15219524 | -0.49318421 | -0.82872182 | -0.4813391 - | -0.50277180 |
| 7 | -0.55664551 | -0.16038760 | 0.09916326 | -0.43369993 | -1.0605580 - | -0.60266908 |
| 8 | 0.02824584 | -0.57997463 | -0.30481495 | 0.02099073 | -0.9320895 - | $-1.08841965$ |
| 9 | 0.52851460 | 0.12503387 | 0.43451870 | -0.74349403 | -1.6961241 - | -1.86023556 |


| year |  |  |
| ---: | ---: | ---: |
| age | 2007 | 2008 |
| 1 | 0.03111028 | -0.05851130 |
| 2 | 0.20874505 | -0.30779484 |
| 3 | -0.25555149 | 0.01175641 |
| 4 | -0.32267868 | -0.07462323 |
| 5 | -0.39252546 | -0.02180249 |
| 6 | -0.52140908 | -0.02610149 |
| 7 | -0.83702182 | 0.28505257 |
| 8 | -1.44675744 | -0.71396474 |
| 9 | -1.35277186 | -0.90728526 |

Table 2.6.2.21 North Sea Autumn Spawning Herring. Fit Parameters

|  | Value | CV.pct | Lower.95.pct.CL |
| :---: | :---: | :---: | :---: |
| F, 2004 | 0.33 | 8 | 0.28 |
| F, 2005 | 0.40 | 9.8 | 0.34 |
| F, 2006 | 0.39 | 9.9 | 0.32 |
| F, 2007 | 0.37 | 10.4 | 0.31 |
| F, 2008 | 0.26 | 8.4 | 0.21 |
| Selectivity at age 0 | 0.18 | 17.6 | 0.10 |
| Selectivity at age 1 | 0.19 | 18.2 | 0.11 |
| Selectivity at age 2 | 0.45 | 11.3 | 0.38 |
| Selectivity at age 3 | 0.71 | 25.8 | 0.60 |
| Selectivity at age 5 | 1.16 | 61.4 | 0.97 |
| Selectivity at age 6 | 1.19 | 56.1 | 0.98 |
| Selectivity at age 7 | 1.06 | 205.6 | 0.85 |
| Terminal year pop, age 0 | 20044856.76 | 1.1 | 13783885.26 |
| Terminal year pop, age 1 | 5971129.00 | 0.9 | 4488422.98 |
| Terminal year pop, age 2 | 3186335.92 | 0.7 | 2590466.26 |
| Terminal year pop, age 3 | 1177819.56 | 0.7 | 973686.34 |
| Terminal year pop, age 4 | 1091076.72 | 0.7 | 907996.60 |
| Terminal year pop, age 5 | 559803.69 | 0.8 | 460826.11 |
| Terminal year pop, age 6 | 508504.34 | 0.8 | 410586.30 |
| Terminal year pop, age 7 | 967084.92 | 0.9 | 751532.58 |
| Terminal year pop, age 8 | 268423.50 | 1.2 | 199834.70 |
| Last true age pop, 2004 | 247136.84 | 1.8 | 160013.05 |
| Last true age pop, 2005 | 246984.42 | 1.4 | 176285.01 |
| Last true age pop, 2006 | 190539.43 | 1.3 | 140911.69 |
| Last true age pop, 2007 | 488487.21 | 1.1 | 367695.30 |
| Recruitment prediction | 32832169.01 | 1.6 | 19169423.75 |
| Index 1, biomass, K | 1.16 | 3.9 | 1.07 |
| Index 1, biomass, Q | 0.00 | 5.2 | 0.00 |
| Index 2, age 0 numbers, Q | 0.00 | 0.5 | 0.00 |
| Index 3, age 1 numbers, Q | 0.00 | 0.7 | 0.00 |
| Index 3, age 2 numbers, Q | 0.00 | 0.9 | 0.00 |
| Index 3, age 3 numbers, Q | 0.00 | 4.5 | 0.00 |
| Index 3, age 4 numbers, Q | 0.00 | 4.3 | 0.00 |
| Index 3, age 5 numbers, Q | 0.00 | 4 | 0.00 |
| Index 4, age 1 numbers, Q | 1.18 | 49.7 | 1.00 |
| Index 4, age 2 numbers, Q | 1.53 | 14.4 | 1.36 |
| Index 4, age 3 numbers, Q | 1.77 | 20.1 | 1.41 |
| Index 4, age 4 numbers, Q | 1.79 | 25.8 | 1.33 |
| Index 4, age 5 numbers, Q | 1.84 | 25.9 | 1.35 |
| Index 4, age 6 numbers, Q | 1.80 | 28.6 | 1.30 |
| Index 4, age 7 numbers, Q | 1.67 | 35.1 | 1.17 |
| Index 4, age 8 numbers, Q | 1.72 | 33.5 | 1.20 |
| Index 4, age 9 numbers, Q | 4.17 | 14.9 | 2.75 |
| SRR, a | 55594838.22 | 1.2 | 36805245.93 |
| SRR, b | 375622.06 | 3.6 | 152326.55 |
|  | Upper.95.pct.CL |  |  |
| F, 20040.39 |  |  |  |
| F, 20050.48 |  |  |  |
| F, 20060.46 |  |  |  |
| F, 20070.46 |  |  |  |
| F, 20080.33 |  |  |  |
| Selectivity at age 0 | 0.32 |  |  |
| Selectivity at age 1 | 0.35 |  |  |
| Selectivity at age 2 | 0.54 |  |  |
| Selectivity at age 3 | 0.84 |  |  |
| Selectivity at age 5 | 1.39 |  |  |
| Selectivity at age 6 | 1.44 |  |  |
| Selectivity at age 7 | 1.32 |  |  |


| Terminal year pop, age 0 | 29149711.77 |  |
| :--- | :--- | :--- |
| Terminal year pop, age 1 | 7943632.26 |  |
| Terminal year pop, age 2 | 3919269.96 |  |
| Terminal year pop, age 3 | 1424749.29 |  |
| Terminal year pop, age 4 | 1311071.44 |  |
| Terminal year pop, age 5 | 680039.97 |  |
| Terminal year pop, age 6 | 629774.21 |  |
| Terminal year pop, age 7 | 1244461.34 |  |
| Terminal year pop, age 8 | 360553.88 |  |
| Last true age pop, 2004 | 381697.75 |  |
| Last true age pop, 2005 | 346037.93 |  |
| Last true age pop, 2006 | 257645.58 |  |
| Last true age pop, 2007 | 648960.56 |  |
| Recruitment prediction | 56232849.54 |  |
| Index 1, biomass, K | 1.25 |  |
| Index 1, biomass, Q | 0.00 |  |
| Index 2, age 0 numbers, $Q$ | 0.00 |  |
| Index 3, age 1 numbers, $Q$ | 0.00 |  |
| Index 3, age 2 numbers, $Q$ | 0.00 |  |
| Index 3, age 3 numbers, $Q$ | 0.00 |  |
| Index 3, age 4 numbers, $Q$ | 0.00 |  |
| Index 3, age 5 numbers, $Q$ | 0.00 |  |
| Index 4, age 1 numbers, $Q$ | 1.38 |  |
| Index 4, age 2 numbers, $Q$ | 1.73 |  |
| Index 4, age 3 numbers, $Q$ | 2.22 |  |
| Index 4, age 4 numbers, $Q$ | 2.40 |  |
| Index 4, age 5 numbers, $Q$ | 2.52 |  |
| Index 4, age 6 numbers, $Q$ | 2.51 |  |
| Index 4, age 7 numbers, $Q$ | 2.38 |  |
| Index 4, age 8 numbers, $Q$ | 2.45 |  |
| Index 4, age 9 numbers, $Q$ | 6.32 |  |
| SRR, a |  | 83976779.91 |
| SRR, b |  | 926246.51 |



Figure 2.6.1.1 North Sea herring. Diagnostics of Acoustic survey catchability at 1 wr from the final ICA assessment. Top left: VPA estimates of numbers at $1 \mathbf{w r}$ (line) and numbers predicted from index abundance at 1 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 1 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 1 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of $\log$ residuals.


Figure 2.6.1.2. North Sea herring. Diagnostics of Acoustic survey catchability at 2 wr from the final ICA assessment. Top left: VPA estimates of numbers at 2 wr (line) and numbers predicted from index abundance at 2 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 2 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 2 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.3. North Sea herring. Diagnostics of Acoustic survey catchability at 3 wr from the final ICA assessment. Top left: VPA estimates of numbers at 3 wr (line) and numbers predicted from index abundance at 3 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 3 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 3 wr . Middle left: $\log$ residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.4. North Sea herring. Diagnostics of Acoustic survey catchability at 4 wr from the final ICA assessment. Top left: VPA estimates of numbers at 4 wr (line) and numbers predicted from index abundance at 4 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 4 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 4 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.5. North Sea herring. Diagnostics of Acoustic survey catchability at 5 wr from the final ICA assessment. Top left: VPA estimates of numbers at 5 wr (line) and numbers predicted from index abundance at 5 wr. Top right: scatterplot of index observations versus VPA estimates of numbers at 5 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 5 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.6. North Sea herring. Diagnostics of Acoustic survey catchability at 6 wr from the final ICA assessment. Top left: VPA estimates of numbers at 6 wr (line) and numbers predicted from index abundance at 6 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 6 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 6 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.7. North Sea herring. Diagnostics of Acoustic survey catchability at 7 wr from the final ICA assessment. Top left: VPA estimates of numbers at 7 wr (line) and numbers predicted from index abundance at 7 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 7 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 7 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.8. North Sea herring. Diagnostics of Acoustic survey catchability at 8 wr from the final ICA assessment. Top left: VPA estimates of numbers at 8 wr (line) and numbers predicted from index abundance at 8 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 8 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 8 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of $\log$ residuals.


Figure 2.6.1.9. North Sea herring. Diagnostics of Acoustic survey catchability at 9+ wr from the final ICA assessment. Top left: VPA estimates of numbers at 9+ wr (line) and numbers predicted from index abundance at $9+$ wr. Top right: scatterplot of index observations versus VPA estimates of numbers at $9+$ wr with the best-fit catchability model (linear function). Middle right: $\log$ residuals of catchability model by VPA estimate of numbers at 9+ wr. Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of $\log$ residuals.


Figure 2.6.1.10. North Sea herring. Diagnostics of IBTS survey catchability at $1 \mathbf{w r}$ from the final ICA assessment. Top left: VPA estimates of numbers at $1 \mathbf{w r}$ (line) and numbers predicted from index abundance at 1 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 1 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 1 wr. Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.11. North Sea herring. Diagnostics of IBTS survey catchability at 2 wr from the final ICA assessment. Top left: VPA estimates of numbers at 2 wr (line) and numbers predicted from index abundance at 2 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 2 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 2 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.12. North Sea herring. Diagnostics of IBTS survey catchability at 3 wr from the final ICA assessment. Top left: VPA estimates of numbers at 3 wr (line) and numbers predicted from index abundance at 3 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 3 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 3 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.13. North Sea herring. Diagnostics of IBTS survey catchability at 4 wr from the final ICA assessment. Top left: VPA estimates of numbers at 4 wr (line) and numbers predicted from index abundance at 4 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 4 wr with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 4 wr . Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.


Figure 2.6.1.14. North Sea herring. Diagnostics of IBTS survey catchability at $5+\mathbf{w r}$ from the final ICA assessment. Top left: VPA estimates of numbers at $5+\mathbf{w r}$ (line) and numbers predicted from index abundance at $5+$ wr. Top right: scatterplot of index observations versus VPA estimates of numbers at $5+\mathbf{w r}$ with the best-fit catchability model (linear function). Middle right: log residuals of catchability model by VPA estimate of numbers at 5+ wr. Middle left: log residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.

## MIK 0-wr, age 0, diagnostics



Figure 2.6.1.15. North Sea herring. Diagnostics of MIK survey catchability at 0 wr from the final ICA assessment. Top left: VPA estimates of numbers at 0 wr (line) and numbers predicted from index abundance at 0 wr . Top right: scatterplot of index observations versus VPA estimates of numbers at 0 wr with the best-fit catchability model (linear function). Middle right: $\log$ residuals of catchability model by VPA estimate of numbers at 0 wr . Middle left: $\log$ residuals of catchability model by year. Bottom left: normal Q-Q plot of $\log$ residuals.

## MLAI, diagnostics



Figure 2.6.1.16. North Sea herring. Diagnostics of MLAI survey catchability at all ages from the final ICA assessment. Top left: VPA estimates of biomass of all ages and biomass predicted from index abundance for all ages. Top right: scatterplot of index observations versus VPA estimates of all ages with the best-fit catchability model (power function). Middle left: log residuals of catchability model by VPA estimate of numbers at 0 wr . Middle right: $\log$ residuals of catchability model by year. Bottom left: normal Q-Q plot of log residuals.

NSH Herring Weighted Residuals Bubble Plot


Figure 2.6.1.17. North Sea herring. Weighted Residuals of surveys and catch for the assessment up to 2009.


Figure 2.6.1.18. North Sea herring. Mean contribution of each indices or catch to the objective function by age.

NSH Herring Retrospective selectivity pattern


Figure 2.6.1.19. North Sea herring. Retrospective selectivity pattern for the year 2000 till 2008.

NSH Herring Retrospective Summary Plot


Figure 2.6.1.20. North Sea herring. Retrospective pattern plots for SSB, Recruits and F2-6


Figure 2.6.1.21. North Sea Herring. Yearclass cohort retrospectives for cohorts that contribute the current stock of North Sea herring.


Figure 2.6.1.22 Model uncertainty; distribution and quantiles of estimated SSB and F2-6 in the terminal year of the assessment. Estimates of precision are based on a parametric bootstrap from the FLICA estimated variance/covariance estimates from the model.

Proportion of Catch.n at age


Figure 2.6.1.23 North Sea Herring. Proportion of catch numbers at age.

Proportion of Catch.wt at age


Figure 2.6.1.24 North Sea Herring. Proportion of catch weight at age.

## Proportion of IBTS index at age



Figure 2.6.1.25 North Sea Herring. Proportion of IBTS index at age.

Proportion of Acoustic index at age


Figure 2.6.1.26 North Sea Herring. Proportion of Acoustic index at age.

## IBTS1: 1-5+ wr



Figure 2.6.1.27 North Sea Herring. Correlation coefficient diagram for IBTS survey.

## Acoustic survey 1-9+ wr



Lower right panels show the Coefficient of Determination $\left(r^{2}\right)$

Figure 2.6.1.2 North Sea Herring. Correlation coefficient diagram for Acoustic survey.


Figure 2.6.1.29 North Sea Herring. Weight at age in the stock over time.
NSH Herring Weight in the stock by cohort


Figure 2.6.1.30 North Sea Herring. Weight at age in the cohort over time.


Figure 2.6.2.1. North Sea herring. Stock summary plot for SSB, recruitment and mean F on ages 26.

## Fitted catch diagnostics



Figure 2.6.2.2. North Sea herring. Diagnostics of selection pattern from the final ICA assessment. Top left: bubbles plot of log catch residuals by age (weighting applied) and year ( 5 yr separable period). Top right: estimated selection parameters (relative to 4 wr ) with $95 \%$ confidence intervals. Bottom left: marginal totals of log residuals by year. Bottom right: marginal totals of log residuals by age (wr).


Figure 2.6.2.3 North Sea Herring. Reference diagrams including indication of reference points assuming a Beverton and Holt stock to recruit relationship. Upper left panel: Equilibrium SSB versus Fishing mortality (ages 2-6). Upper right panel: Recruit versus SSB relationship. Bottom right panel: Yield versus Fishing mortality (ages 2-6). Bottom right: Yield versus SSB. Grey points indicate Bmš and Fmš


Figure 2.6.2.4. North Sea herring. Agreed management plan for adult fishery (A-fleet, ages 2-6) including trigger biomass points ( $B_{l i m}$ and $B_{\text {trigger }}$ ) and $B_{\text {pa. }}$. Black dots represent realised estimated fishing mortalities from 2002 untill 2008. Fishing mortality in 2009 is estimated based on the agreed TACS for the A-fleet from the short term prediction (see section 2.7).


Figure 2.6.2.5. North Sea herring. Stock and recruit plot. Each point labelled by year class.

### 2.7 Short term predictions

Short term predictions for 2010 were done with the MFSP software, following the procedures in the stock annex 3 . This assumes that recruitment will continue to be at the low level since 2002 (geometric mean of 2001 to 2007 year classes)

For the intermediate year, an overshoot of $13 \%$, which is the average overshoot in 2006 - 2008 for the A fleet was assumed. For the B, C and D fleets the same fraction of the TAC as last year was assumed. See table 2.7.1 for other inputs.

The seven scenarios presented below are based on an interpretation of the harvest control rule or other options and are only illustrative:
a) No fishing;
b) The EU-Norway management plan;
c) A roll over TAC from 2009 to 2010 of 171 kt for the A fleet;
d) Catches that are estimated lead to SSB $>$ Bpa in 2011;
e) A $15 \%$ decrease in A fleet in TAC between 2009 and 2010;
f) A 15\% increase in A fleet in TAC between 2009 and 2010;
g) Option b but with larger catches of approximately $40 \%$ for the $C$ and D fleet

Since the current management plan only stipulates overall fishing mortalities for juveniles and adults, making fleet-wise predictions for four fleets that are more or less independent provides different options for 2010. The consequence of other combinations of catch options can be explored on request.

For options b, c, e and f, the C and D fleets are assumed to have a catch for 2010 of 39800 tonnes, giving expected catches of 7.4 and 3.7 thousand tonnes respectively of North Sea autumn spawners. For option g the catches of the C and D fleet are assumed to be 56600 tonnes.

All predictions are for North Sea autumn spawning herring only. The results are presented in Table 2.7.2.

### 2.7.1 Comments on the short-term projections

HAWG assumed that recruitment was likely to remain poor in 2010. A slight decrease in SSB is expected from 2008 to 2009 . The SSB is expected to increase slightly both in 2010 and further in 2011, indicating that the current management has the potential to reverse the decline in the stock and stabilize it above the present level. The SSB is not expected to reach Bpa in 2010 even without fishery, but it may reach Bpa in 2011 with a substantial reduction (well over 30\%) in catches.

The estimated impact of the juvenile fishery depends on the assumed value for natural mortality. It has not been investigated to what extent changes in natural mortality would affect the current advice, or if indeed such changes are taking place. Some of the important predator stocks are currently in a poor condition.

The predictions this year are in line with those obtained last year. The predicted catch according to the harvest rule implies less reduction than $15 \%$.

Table 2.7.1. North Sea autumn spawning herring. The input file used for the short term prediction of North Sea herring.



Table 2.7.2. North Sea autumn spawning herring. Management options for North Sea herring.
Outlook assuming a TAC constraint for fleet A in 2009
Basis: Intermediate year (2009) with catch constraint

| F <br> fleet <br> A | F <br> fleet <br> B | F <br> fleet <br> C | F <br> fleet <br> D | Fo-1 $^{\text {F2-6 }}$ | Catch <br> fleet A | Catch <br> fleet B | Catch <br> Fleet C | Catch <br> fleet D | SSB <br> 2009 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.184 | 0.021 | 0.006 | 0.004 | 0.04 | 0.189 | 194.2 | 7.4 | 6.5 | 2.7 | 971 |

Scenarios for prediction year (2010)

|  | F-values by fleet and total |  |  |  |  |  | Catches by fleet |  |  |  | results |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | F <br> Fleet A | F <br> FLEET B | F <br> FLEET C | $\mathrm{F}$ <br> FLEET D | $\mathrm{F}_{0-1}$ | $\mathrm{F}_{2-6}$ | CATCH <br> FLEET <br> A | CATCH <br> FLEET <br> B | CATCH <br> FLEET <br> C | CATCH <br> FLEET <br> D | $\begin{aligned} & \hline \text { SSB } \\ & 2010^{1)} \end{aligned}$ | $\begin{aligned} & \hline \text { SSB } \\ & 2011 \end{aligned}$ | $\begin{aligned} & \text { \%SSB } \\ & \text { change } \\ & \text { 2) } \end{aligned}$ | \%TAC <br> change <br> fleet A <br> 3) |
| a | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1133 | 1497 | 17\% | -100\% |
| b | 0.144 | 0.033 | 0.005 | 0.006 | 0.050 | 0.148 | 164.3 | 10.4 | 7.4 | 3.7 | 1027 | 1209 | 6\% | -4\% |
| c | 0.150 | 0.033 | 0.005 | 0.006 | 0.050 | 0.155 | 171.0 | 10.4 | 7.4 | 3.7 | 1023 | 1198 | 6\% | 0\% |
| d | 0.094 | 0.022 | 0.003 | 0.004 | 0.033 | 0.097 | 110.2 | 7.0 | 5.0 | 2.5 | 1063 | 1300 | 9\% | -36\% |
| e | 0.126 | 0.033 | 0.005 | 0.006 | 0.049 | 0.131 | 145.4 | 10.4 | 7.4 | 3.7 | 1040 | 1238 | 7\% | -15\% |
| f | 0.174 | 0.033 | 0.005 | 0.006 | 0.051 | 0.179 | 196.7 | 10.4 | 7.4 | 3.7 | 1007 | 1160 | 4\% | 15\% |
| g | 0.144 | 0.028 | 0.007 | 0.008 | 0.050 | 0.149 | 162.6 | 9.0 | 10.5 | 5.3 | 1027 | 1206 | 6\% | -5\% |

Weights in ' 000 t .
Shaded areas are considered not in accordance with the precautionary approach.
All numbers apply to North Sea autumn-spawning herring only.
${ }^{1)}$ For autumn spawning stocks, the SSB is determined at spawning time and is influenced by fisheries between $1^{\text {st }}$ January and spawning.
${ }^{2)}$ SSB (2010) relative to SSB (2009).
${ }^{3}$ Calculated landings (2009) relative to TAC 2008.

### 2.8 Medium term predictions and HCR simulations

The ICES workshop on herring management plans (WKHMP, ICES CM 2008 (ACOM:27)) met in February 2008 and carried out extensive investigation of the medium term scenarios for North Sea herring (see section 1.3), this lead to an adjustment of the management plan in November 2008. Further analysis was thus not carried out by HAWG.

### 2.9 Precautionary and Limit Reference Points

The precautionary reference points for this stock were adopted in 1998. The situation has now arisen that North Sea herring is nominally being managed by a precautionary management plan, although the SSB is now below the precautionary biomass reference point. We consider that the critical issue is identifying the risk of SSB falling below Blim. The following section is adapted from ICES WKHMP (ICES CM 2008 (ACOM:27)) and explores and discusses the issues about precautionary status of the management of North Sea herring.

## The Blim

The 1998 Study Group on Precautionary Approach to Fisheries Management (ICES CM 1998/ACFM:10.) determined reference points for North Sea herring that were adopted by ICES (ICES CM 1998/ACFM:10.). The Blim (800 000 tonnes) was set at a level below which the recruitment may become impaired and was also the formally used MBAL. In 2007, WKREF (ICES CM 2007/ACFM:05) explored limit reference points for North Sea herring and concluded that there is no basis for changing Blim. A low risk of SSB falling below Blim is therefore the basis of ICES precautionary advice.

## Fpa and Bpa

The target and trigger points used in the management plan (which began in 1997) were recommended by the Study Group on Precautionary Approach to Fisheries Management and adopted by ICES as the precautionary reference points. This means that the precautionary reference points were taken from the already existing management plan. In the management plan, the target fishing mortalities were intended as targets and not as bounds. The higher inflection point (B trigger) in the earlier rule ( 1.3 million tonnes) was derived largely as a compromise, allowing higher exploitation at higher biomass but reflecting an ambition to maintain the stock at a high level, by reducing the fishing mortality at an early stage of decline. This trigger was changed in November 2008 to 1.5 million tonnes after WKHMP and consultation with the stakeholders. Thus currently the trigger and Bpa are different at 1.5 million tonnes and 1.3 million tonnes respectively.

## Concept of a management plan (harvest control rule)

In a harvest control rule, parameters (trigger and targets) serve as guidance to actions according to the state of the stock (ICES Study Group on the Precautionary Approach, ICES CM 2002/ACFM:10). These should be chosen according to management objectives, one of which should be to have a low risk of bringing the SSB to unacceptably low levels. In the evaluation of a harvest rule, one will use simulations with a 'virtual stock' which as far as possible resembles the stock in question, and the risk is evaluated as the probability of the virtual SSB being below the Blim value. Within the constraints needed to keep the risk to Blim low, parameters of the rule will be
chosen to serve other management objectives, e.g. to ensure a high long term yield and stable catches over time. Such a management plan would be classed by ICES as precautionary provided the risk of SSB being below Blim is sufficiently low. This conforms to the recommendations of WKOMSE (ICES 2009).

## Concept of precautionary reference points

Conceptually, precautionary reference points (Bpa) are different from parameters in a harvest control rule. In the precautionary approach, as interpreted by ICES, the function of the reference points is to ensure that the SSB is above the range where recruitment may be impaired or the stock dynamics is unknown. The real limit is represented by Blim, while the Bpa takes assessment uncertainty into account, so that if SSB is estimated at Bpa, the probability that it is below Blim shall be small. The Flim is the fishing mortality that corresponds to Blim in a deterministic equilibrium. The Fpa is related to Flim the same way as Bpa is related to Blim (ICES Study Group on the Precautionary Approach 2002b). In the advisory practice, Fpa has been the basis for the advice unless the SSB has been below Bpa, where a reduction in F has been advised. Furthermore, Fpa and Bpa are currently used to classify the state of stock and rate of exploitation relative to precautionary limits. Precautionary reference points are used by ICES to provide advice and classify the state of the stock in the absence of other information, such as extensive evaluations of management plans.

## Conclusion

ICES will accept that a harvest control rule is in accordance with the precautionary approach as long as it implies a low risk to being below Blim, even if other reference points may be exceeded occasionally. When a rule is regarded as precautionary, ICES gives its advice according to the rule. If the rule is followed, then ICES classifies exploitation as precautionary. Within this framework, other precautionary reference points generally will be redundant. However, the precautionary reference points may also be used to classify the stock with respect to precautionary limits, which may lead to a conflicting classification. This discrepancy is still unresolved. For North Sea herring in the present situation, with a reduced recruitment, the SSB may be expected to be below 1.3 million tonnes most of the time. The management plan will reduce fishing mortality accordingly. Following the acceptance by ICES that the management plan is precautionary (and the findings of WKHMP), HAWG considers that the parameters of the management plan should take primacy over the management against precautionary reference points Fpa or Bpa.

### 2.10 Quality of the Assessment

### 2.10.1 Precision of the estimates

The precision of the assessment derived from the FLICA model is based on a parametric variance covariance bootstrap of the parameters that influence the estimates of terminal F and SSB. The estimated precision expressed as a percentile contour plot is shown in Figure 2.6.1.22. The $95 \%$ intervals are given for F and SSB in combination and separately.

### 2.10.2 Comparison with earlier assessments

The historical evaluation of the NS herring assessment from 1990 to 2008 Simmonds (2009) supports the contention of a precise assessment, particularly in its current configuration since 2002. This years diagnostics continue to support that view. An eight year analytic retrospective (Figure 2.6.1.20) shows the current consistency of the assessment. The data from the stock summary table is compared with the stock summary from the 2008 assessment and the first year (intermediate year) of the 2008 short term prediction. With the exception of the estimate of recruitment age 0 in 2008 the 2009 assessment is in good agreement with the assessment carried out last year, see text table below.

|  | 2008 AsSESSMENT |  |  |  | 2009 AsSESSMENT |  |  |  | Percentage Change in ESTIMATE 2008-2009 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Rec | SSB | Catch | $\mathrm{F}_{2}$-6 | Rec | SSB | Catch | $\mathrm{F}_{2-6}$ | Rec | SSB | Catch | $\mathrm{F}_{2-6}$ |
| 2006 | 25024 | 1252 | NA | 0.35 | 27695 | 1234 | NA | 0.35 | 11\% | -1\% | NA | 0\% |
| 2007 | 20853 | 977 | NA | 0.33 | 19044 | 953 | NA | 0.34 | -9\% | -2\% | NA | 3\% |
| 2008 | 9223 | 978 | 240 | 0.21 | 22909 | 1000 | 258 | 0.24 | 148\% | 2\% | 8\% | 14\% |

* projected values from the intermediate year in the deterministic short term projection, assuming catch constraint with small overshoot. (Recruits are defined as age 0 )

The revision on F and SSB are all small. The revision of recruitment at $+148 \%$ is more important. In 2008 the recruitment (age 0 ) was only estimated by one survey (MIK), this year (in 2009) that cohort has additional estimates from 1 year of catch and the IBTS survey. To try to address this issue adapted time-series of MIK 0wr and IBTS 1wr have been calculated to take into account more biological realism (Table 2.3.3.2). These account for the presence of the Downs herring in the MIK survey and the presence of some 1wr WBSS herring in the IBTS in the Kattegat. Fitting these adapted indices in ICA assessment very slightly improves the fit in the model. The changes are mostly coming from the fit to the IBTS 1 wr , where 19 of the 26 residuals are reduced in magnitude. In contrast although the MIK fits better overall, there are increases and decreases in the magnitude of residuals in almost equal numbers of years. The initial indications are that SSB would be estimated as 3\% lower and F 3\% higher in 2008. These adapted series need more exploration before they can be used in an assessment, but they do not indicate substantive revision to terminal values.

The cohort retrospective evaluations suggest the WG is providing a very consistent evaluation of most year classes (Figure 2.6.1.21). The exceptions are the 2001, 2004 and 2006 year classes which are more variable in the first two years of observations. In particular the large 2000 year class has been estimated consistently since it was first seen in 2001.

The both assessment and projections currently appear to be a good basis for management advice.

### 2.11 Herring in Division IVc and VIId (Downs Herring).

Over many years the working group has attempted to assess the contribution of winter spawning Downs herring to the overall population of North Sea herring. Since 1985, there is a separate TAC for herring in Divisions IVc and VIId as part of the total North Sea TAC.

Historically, the TAC for herring in IVc and VIId has been set as a proportion of the total North Sea TAC and this has varied between 6 and $16 \%$ since 1986. The proportion has been relatively high, particularly between 2002 and 2005. However, ICES in 2005 expressed concerns regarding Downs herring and recommended that the proportion used to determine the TAC should be set to the long term average of the proportions used since 1986 (11\%). In accordance with ICES advice the sub-TAC was cut by $33 \%$ in 2006 and the proportion was kept to $11 \%$ of the human consumption TAC in 2007. For 2008, it was set at 26771 tonnes and at 23567 tonnes for 2009, which represents respectively $13 \%$ and $14 \%$ of the total human consumption TAC for Divisions IV and VIId (Figure 2.11.1).

ICES has in the past expressed concern that there is a persistent tendency to overfish the Downs TAC. However, this tendency has been markedly reduced in recent years, mainly for the two last years (Figure 2.11.2). Landings in 2008 amounted to 29600 tonnes ( 39000 tonnes in 2007).

Historically, the Downs herring has been considered highly sensitive to overexploitation (Burd, 1985; Cushing 1968; 1992). It is less fecund and expresses different growth dynamics and recruitment patterns to the more northern spawning components. Furthermore, the directed fishery in Q4 and Q1 targets aggregations of spawning herring. Preliminary studies undertaken by HAWG in 2006 (ICES CM 2006/ACFM:20) based on population profiles suggested that total mortality ( $Z$ ) was significantly higher for the 1998 and 1999 year classes of Downs herring compared to herring caught in the northern part of the North Sea.

Downs herring is also taken in other herring fisheries in the North Sea. Downs herring mixes with other components of North Sea herring in the summer whilst feeding. Analysis of Dutch catches from this summer fishery, suggests that in recent years equal proportions of autumn and winter (Downs) herring were caught (Figure 2.11.3, see WD 10 for methods). There is also a summer industrial fishery in the eastern North Sea exploiting Downs and North Sea autumn spawning herring juveniles. Tagging experiments in the Eastern North Sea (Aasen et al, 1962) estimated that around $15 \%$ of those catches comprised Downs recruits. Otolith microstructure studies of catches from the northern North Sea suggested that the proportion of Downs herring may vary considerably from year to year ( 26 to $60 \%$ ) and may also vary between fleets.

The proportion of the autumn and winter spawning components in recruiting year classes of North Sea herring has been traditionally monitored through the abundance of different sized fish in the IBTS. The 1-ring fish from Downs spawning sites (winter) are thought to be smaller ( $<13 \mathrm{~cm}$ ) than those from the more northern, autumn spawning sites $(>13 \mathrm{~cm})$. Both the total abundance and the proportion of Downs herring have, on average, been comparatively higher since the early 1990s, although there is considerable variation between year classes (Table 2.3.3.3 and Figure 2.11.4). These size data suggest that around $35 \%$ of the strong 2000 year class came from Downs production and approximately $70 \%$ of the 2002 year class (Figure 2.11.4). This is support by the analysis of the summer catches (WD10). For the 2007 year class, the
percentage of $27 \%$ corresponds to the mean value of the time serie. However the number of small herrings shows signs of increasing since 2004 (Figure 2.11.5).

Using a ten layered finite-volume advection-dispersion model with real time meteorological and freshwater runoff drivers, the interannual variability in transport of Downs herring larvae transport, appears to be linked to recruitment variability (Dickey-Collas et al., WD4.) as shown. Almost all Downs herring larvae move east after hatching and simulations suggest that meteorologically forced transport will deliver the Downs herring larvae to the nursery grounds in the south-eastern North Sea. However it is not the delivery of the larvae that appears to be positively related to year class strength but rather the retention.

As mentioned on section 2.3.3.1 the MIK hauls for 0-ringers in this area also include Downs herring larvae. However, at the time of the IBTS survey these larvae are relatively small compared to herring larvae from other stocks. Accordingly their accumulated mortality to recruitment will be relatively higher compared to the larvae from the other stocks. Therefore the small larvae ( $<20 \mathrm{~mm}$ ) have until now been excluded from the standard estimation of 0-ringer recruitment. During the present meeting of the WG, trials have been made to investigate the possibilities and consequences of including these in recruitment estimation for the North Sea herring stock. This is further described in section 2.10.2

In 2007, the extension of the IBTS $1^{\text {st }}$ quarter survey area in the Eastern English Channel was implemented in the survey design: additional GOV hauls and MIK stations carried out in this area have provided more information on Downs herring. (ICES CM 2007/ACFM:11). This sampling continued in 2008 and 2009.

Acoustic data recorded at the same time (January 2009) and in the same way as previous years showed that important herring schools were still along the French coast at this time of the years. The catch composition of the pelagic hauls consisted of herring smaller than previous years of 23 cm mean length fish ( 27.5 cm in 2008) belonging to age-groups 2-3. (Figure 2.11.6). Large and continuous shoals of herring were found at the survey time in a localised area, which the mean density could be estimated around 1500 tonnes per nautical mile square but it could not be raised to the whole area due to the spatial heterogeneity and the sampling protocol used.

In conclusion, the TAC is specific to the conservation of the spawning aggregation of Downs herring. In the absence of other information there are uncertainties in the recruitment to the component in the next few years and HAWG recommends that the IVc-VIId TAC should be maintained at $11 \%$ of the total North Sea TAC (as recommended by ICES). This recommendation should be seen as an interim measure prior to the development of a more robust harvest control rule for setting the TAC of Downs herring, supported by increased research effort into the dynamics of this component in fisheries in the central and northern North Sea. Any new approach should provide an appropriate balance of F across stock components and be similarly conservative until the uncertainty in the Downs contribution to the catch in all fisheries in the North Sea is reduced. Methods illustrated by Kell et al. (2009) may be appropriate.


Figure 2.11.1. North Sea herring. Comparison of TACs for total North Sea and IVc and VIId


Figure 2.11.2. Downs herring in IVc and VIId. Comparison of historical catches and TACs


Figure 2.11.3 North Sea herring. Estimated totals landed catch by spawner types from the Dutch summer (May to July fishery). Open circles: autumn spawners; open triangles: winter spawners. Error bars denote standard errors.


Figure 2.11.4. Downs herring. Proportion of small 1-ringers versus all sizes in the North sea (from table 2.3.3.3)


Figure 2.11.5. Downs herring. Index (Nos per hr) of small ( $<13 \mathrm{~cm}$ ) 1-ringers in the North from table 2.3.3.3).


Figure 2.11.6. Downs herring. IBTS 09. Catch composition by age from pelagics hauls in the Eastern English Channel. The total percentage for the 2,3 and 4 ringers represent respectively $38 \%$, $30 \%$ and $10 \%$.

### 2.12 Management Considerations

Based on the most recent estimates of SSB and fishing mortality, ICES classifies the stock as being at risk of having reduced reproductive capacity and is beingharvested sustainably. The SSB in autumn 2008 was estimated at 1.0 million t , and is expected to remain below Bpa ( 1.3 million t ) in 2009. $\mathrm{F}_{2-6}$ in 2008 was estimated at 0.24 , above the target $\mathrm{F}_{2-6}$ of 0.14 . The year classes since 2002 are estimated to be among the weakest since the late 1970s.

The stock is managed according to the EU-Norway Management agreement which was updated on November 2008 (Table 2.12.1).

WKHMP examined the performance of this management plan and the plan is consistent with the precautionary approach because of the low risk of $\mathrm{SSB}<\mathrm{Blim}$.

SSB and fishing mortality are reliably estimated. A reduction in fishing mortality to close to the target is expected to be achieved in 2009. The 2008 year class is estimated within the range of low recruitment. Therefore HAWG assumes that the recruitment will remain at the lower level. Delay in implementing substantial reductions in catch by not following the management plan has resulted in the SSB being at greater risk of being below Blim and in lower catches. The management plan should be followed.

North Sea herring and Western Baltic Spring Spawning herring are managed under mixed quotas in some areas of North Sea, Skagerrak and Kattegat. The management of these mixed components was discussed in detail in 2007 (ICES CM 2007 ACFM:11). With the decline of both the WBSS herring and the NS herring, conservation of both stock needs to be considered when setting TACs. With the mixing of stocks within a fishery, primacy of consideration should be given to protection of the stock most heavily exploited in the area of overlap.

The options selected for the C- and D-fleets are compatible with the advised exploitation of Western Baltic Spring Spawners assuming a TAC for 2010 of 39800 tonnes (see Section 6.4.7) and are 7.4 and 3.7 thousand tonnes of North Sea autumn spawning herring for $C$ and $D$ fleets respectively. A further option assuming higher catches of the C and D fleet (approximately $40 \%$ higher) results in lower catches for the A and B fleet.

The North Sea autumn spawning herring stock also includes the Downs herring component (herring in Divisions IVc and VIId), the management of this component was discussed in detail in 2007 (ICES CM 2007 ACFM:11). There is no update to this advice.

Table 2.12.1

According to the EU-Norway agreement (November 2008):

1. Every effort shall be made to maintain a minimum level of Spawning Stock Biomass (SSB) greater than 800,000 tonnes (Blim).
2. Where the SSB is estimated to be above 1.5 million tonnes the Parties agree to set quotas for the directed fishery and for bycatches in other fisheries, reflecting a fishing mortality rate of no more than 0.25 for 2 ringers and older and no more than 0.05 for $0-1$ ringers.
3. Where the SSB is estimated to be below 1.5 million tonnes but above 800,000 tonnes, the Parties agree to set quotas for the direct fishery and for bycatches in other fisheries, reflecting a fishing mortality rate on 2 ringers and older equal to:
$0.25-\left(0.15^{*}(1,500,000-S S B) / 700,000\right)$ for 2 ringers and older, and no more than 0.05 for $0-1$ ringers
4. Where the SSB is estimated to be below 800,000 tonnes the Parties agree to set quotas for the directed fishery and for bycatches in other fisheries, reflecting a fishing mortality rate of less than 0.1 for 2 ringers and older and of less than 0.04 for $0-1$ ringers.
5. Where the rules in paragraphs 2 and 3 would lead to a TAC which deviates by more than $15 \%$ from the TAC of the preceding year the parties shall fix a TAC that is no more than $15 \%$ greater or $15 \%$ less than the TAC of the preceding year.
6. Notwithstanding paragraph 5 the Parties may, where considered appropriate, reduce the TAC by more than $15 \%$ compared to the TAC of the preceding year.
7. Bycatches of herring may only be landed in ports where adequate sampling schemes to effectively monitor the landings have been set up. All catches landed shall be deducted from the respective quotas set, and the fisheries shall be stopped immediately in the event that the quotas are exhausted.
8. The allocation of the TAC for the directed fishery for herring shall be $29 \%$ to Norway and $71 \%$ to the Community. The bycatch quota for herring shall be allocated to the Community.
9. A review of this arrangement shall take place no later than 31 December 2011.
10. This arrangement enters into force on 1 January 2009.

### 2.13 Ecosystem considerations

### 2.13.1 Ecosystem considerations

Herring is considered to have a major impact on most other fish stocks as prey and predator and is itself prey for seabirds and sea mammals in that area. Herring spawning and nursery areas, being near the coasts, are particularly sensitive and vulnerable to anthropogenic influences. The most serious of these is the extraction of marine sand and gravel and the development of coastal wind farms. Herring leave and then repopulate spawning grounds and the lack of spawning in recent years does not mean that the spawning ground is not required to maintain a resilient herring population.

The human consumption fisheries for herring are considered relatively clean, with little by-catch of other fish and almost no disturbance of the sea bed. The limited evidence from observer programmes suggest that discarding of herring is not widespread. Juvenile herring are caught as a by catch of industrial fisheries and these vessels catch a range of fish species. There is little information available on the catches of mega-fauna by the herring fleets.

### 2.13.2 Changes in the environment

This stock has recently produced six poor year classes in a row, which has never been observed before. Larval surveys show a large abundance of larvae in recent years. However, survival of these larvae seems to be very poor. The specific reasons for this are not known. An ICES study group has reviewed the hypotheses for the serial poor recruitment in North Sea herring (Payne et al., 2009) and commented that the reduction in herring recruitment is similar to the warming of the water on the spawning grounds and changes in the hydrography. These hydrographic changes may be linked to the AMO (Atlantic Multidecadal Oscillation) and are also associated with changes in the zooplankton community. Further investigation of the causes of the poor recruitment will require targeted research projects.

## 3 Herring in Division IIIa and Subdivisions 22-24 [update assessment]

### 3.1 The Fishery

### 3.1.1 Advice and management applicable to 2008 and 2009

A benchmark assessment was carried out in 2008. SSB appeared stable over a number of years. Fishing mortality estimates for 2007 were 0.47 for adults (Fbar) and 0.17 for the juveniles (1-ringers). The recruitment demonstrated a declining trend since 2003 and fishing mortality was estimated at a stable high level compared to other herring stocks. Because of the very low recruitment in recent years ACOM recommended in 2008 a substantial reduction in fishing mortality in 2009. In the absence of a management plan and agreed target and precautionary reference points ICES advised that fishing mortality should be less than the F related to high long-term yield ( $\mathrm{F}=0.25$ ). This would correspond to landings of less than 32800 t in 2009.

The EU and Norway agreement on a herring TAC for 2008 was 51673 t in Division IIIa for the human consumption fleet and a by-catch ceiling of 11470 t to be taken in the small mesh fishery. For 2009, the EU and Norway agreement on herring TACs in Division IIIa was 37722 t for the human consumption fleet and a by-catch ceiling of 8 373 t to be taken in the small mesh fishery..

Previous to 2006 no special TAC for Subdivisions 22-24 was set. In 2008, a TAC (44 550 t ) was set on the Western Baltic stock component. The TAC for 2009 was set at 27 176 t .

### 3.1.2 Catches in 2008

Herring caught in Division IIIa are a mixture of North Sea Autumn Spawners (NSAS) and Western Baltic Spring Spawners (WBSS). This Section gives the landings of both NSAS and WBSS, but the stock assessment applies only to the spring spawners.

Landings from 1987 to 2008 are given in Table 3.1.1 and Figure 3.1.1. In 2008 the total landings in Division IIIa and Subdivisions 22-24 have decreased to 81300 t , which is the lowest value of the time series (1986-2008). The decrease in landings in 2008 is particularly evident in the Kattegat, where the Swedish landings were less than half of the landings in 2007 from that area. The German landings have increased slightly for the last three years in Subdivision 22-24, but are still small in Division IIIa. As in previous years the 2008 landing data are calculated by fleet according to the fleet definitions used when setting TACs.

The fleet definitions used since 1998 are:
Fleet C: directed fishery for herring in which trawlers (with 32 mm minimum mesh size) and purse seiners participate.

Fleet D: All fisheries in which trawlers (with mesh sizes less than 32 mm ) and small purse seiners, fishing for sprat along the Swedish coast and in the Swedish fjords, participate. For most of the landings taken by this fleet, herring is landed as by-catch. Danish and Swedish by-catches of herring from the sprat fishery and the Norway pout and blue whiting fisheries are listed under Fleet D.

Fleet F: Landings from Subdivisions 22-24. Most of the catches are taken in a directed fishery for herring and some as by-catch in a directed sprat fishery.

In Table 3.1.2 the landings are given for 2002 to 2008 in thousands of tonnes by fleet (as defined by HAWG) and quarter.

Selection by fleet is done disregarding the nationality of the fleets assuming that the fleets target the same part of the population regardless of national flag. However, analysing of the age distribution in the catches of the Danish and Swedish Fleet D in Subdivisions 20 and 21 it became apparent that the Swedish Fleet D targets a larger part of the population as the landings of fish older than 3 years are higher than what is observed in the Danish catches of the same fleet. Thus the selection by fleet is not identical between the two countries. The Danish fleet definition follows the definition set by HAWG, where Fleet D (or so called Industrial fleet) is defined as all fisheries in which trawlers (with mesh sizes less than 32 mm ) and small purse seiners, fish for sprat. For most of the landings taken by this fleet, herring is landed as by-catch from the sprat fishery and the Norway pout and blue whiting fisheries. The Swedish fleet definition is based on mesh size of the gear, as for the Danish fleet. However, an earlier change in the Swedish industrial fishery implies that there is no difference in age structure of the landings between vessels using different mesh sizes since both are basically targeting herring for human consumption. Thus Swedish age-length keys cannot be used to raise Danish catches and vice versa.

The text table below give the TACs and Quotas ( t ) for the fishery by the C- and Dfleets in Division IIIa and for the F-fleet in Subdivisions 22-24.

|  | TAC | DK | GER | SF | PL | SWE | EC | NOR | FAROE |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 2008 |  |  |  |  |  |  |  |  |
| Div. IIIa fleet-C | 51,673 | 21,474 | 344 |  |  | 22,463 | 44,281 | 6,892 | 500 |
| Div. IIIa fleet-D | 11,470 | 9,805 | 87 |  |  | 1,578 | 11,470 |  |  |
| SD 22-24 fleet-F | 44,550 | 6,245 | 24,579 | 3 | 5,797 | 7,926 | 44,550 |  |  |
| \% of IIIa taken in IV |  |  |  |  |  |  |  | $-30 \%$ |  |
|  | 2009 |  |  |  |  |  |  |  |  |
| Div. IIIa fleet-C | 37,722 | 15,611 | 250 |  |  | 16,329 | 32,190 | 5,032 | 500 |
| Div. IIIa fleet-D | 8,373 | 7,157 | 64 |  |  | 1,152 | 8,373 |  |  |
| SD 22-24 fleet-F | 27,176 | 3,809 | 14,994 | 2 | 3,536 | 4,835 | 27,176 |  |  |
| \% of IIIa taken in IV |  |  |  |  |  |  |  | $-20 \%$ |  |

### 3.1.3 Regulations and their effects

In recent years, HAWG has calculated a substantial part of the catch reported as taken in Division IIIa in fleet C actually has been taken in Subarea IV. These catches have been allocated to the North Sea stock and accounted under the A-fleet. Estimates based on VMS and Industry information suggest that $36 \%, 28 \%$ and $30 \%$ of the official landings for human consumption in Division IIIa have been misreported in the last three years, respectively. These figures are probably underestimating the problem since only a subset of countries supply this information to the HAWG. Misreported catches have been moved to the appropriate stock for the assessment.
Regulations allowing quota transfers from Division IIIa to the North Sea were introduced as an incentive to decrease misreporting for the Norwegian part of the fishery, the percentage has gradually been decreased in recent years being $20 \%$ in 2009.

The quota for the $C$ fleet and the by-catch quota for the $D$ fleet (see above) are set for the NSAS and the WBSS stocks together. The implication for the catch of NSAS must also be taken into account when setting quotas for the fleets that exploit these stocks.

### 3.1.3.1 Changes in fishing technology and fishing patterns

There have been no significant changes in fishery technology in the last few years.

### 3.2 Biological composition of the catch

Table 3.2.1 and Table 3.2.2 show the total catch (autumn- and spring-spawners combined) in numbers and mean weight-at-age in the catch for herring by quarter and fleet landed from Skagerrak and Kattegat, respectively. The total catch in numbers and mean weights-at-age for herring landed from Subdivisions 22-24 are shown in Table 3.2.3.

The level of sampling of the commercial landings was generally acceptable (Table 3.2.4). In the cases of missing samples the corresponding landings were minor. Where sampling was missing in areas and quarters on national landings, sampling from either other nations or adjacent areas and quarters were used to estimate catch in numbers and mean weight-at-age (Table 3.2.5).

Based on the proportions of spring- and autumn-spawners in the landings, catches were split between NSAS and WBSS (Table 3.2.6 and see Section 3.2.2 for more details)

The total numbers and mean weight-at-age of the WBSS and NSAS landed from Kattegat, Skagerrak, and Division IIIa respectively was then estimated by quarter and fleet (Table 3.2.7-3.2.12).

The total catch, expressed as SOP, of the WBSS taken in the North Sea + Div. IIIa in 2008 was estimated to be 25200 t , and has thereby decreased to below the levels observed in 2003 ( 38000 t ) and 2004 ( 35000 t ) from the somewhat high level in 2006 (48 700 t) (Table 3.2.13).

Total catches of WBSS from the North Sea, Division IIIa, and Subdivisions 22-24 respectively, by quarter, was estimated for 2008 (Table 3.2.14). Additionally, the total catches of WBSS in numbers and tonnes, divided between the North Sea and Division IIIa and Subdivisions 22-24 respectively for 1992-2008, are presented in Tables 3.2.15 and 3.2.16.

Catches of WBSS from Subdivisions 22-24 have remained rather stable since 2003. In 2008 the catch was again around the same level at 43000 t (Table 3.2.16).

The total catch of NSAS in Div. IIIa amounted to 12949 t in 2008, which is the lowest value observed in the time series (Table 3.2.17).

### 3.2.1 Quality of Catch Data and Biological Sampling Data

No quantitative estimates of discards were available to the Working Group. However, the amount of discards for 2008 is assumed to be insignificant, as in previous years.

Table 3.2.4 shows the number of fish aged by country, area, fishery and quarter. The overall sampling in 2008 more than meets the recommended level of one sample per 1000 t landed per quarter and the coverage of areas, times of the year and gear (mesh size) was acceptable. One exception is the scarce sampling covering catches from Subdivision 23 comprising $5660 t$, where the amount of total samples match the recommended level, but the temporal coverage is not acceptable. However, for some of the sampling units (SD and quarter) sample size of ages was possibly lower (see HAWG WD Cardinale and Hansson 2006 for details) than the value necessary to
reach $\pm 5 \%$ precision level as established by the current European Data Regulation system.

### 3.3 Fishery Independent Information

### 3.3.1 German Acoustic Survey in Subdivisions 21-24 (Autumn)

A joint German-Danish acoustic survey was carried out with R/V "SOLEA" between 2 and 21 October 2008 in the Western Baltic covering Subdivisions 21, 22, 23 and 24. A full survey report is given in the Report of the Planning Group for Herring Surveys (ICES 2009/LRC:02). The results for 2008 are presented in Table 3.3.1. The time series has been revised in 2008 (ICES 2008/ACOM:62) to include the southern part SD 21. The years 1991-1993 were excluded due to different recording method at that time and 2001 was also excluded since SD 23 was not covered (ICES 2008/ACOM:62). The Western Baltic spring spawning herring stock in 2008 was estimated to be $3.4 \times 10^{9}$ fish or about $118 \times 10^{3}$ tonnes in Subdivisions 21-24. Those estimates are comparable to levels of abundance and biomass observed in 2007 (Table 3.3.1).

### 3.3.2 Herring Acoustic Survey (HERAS) in Division Illa (Summer)

The Herring acoustic survey (HERAS) from 26 June to 10 July 2008 covered the area in the Skagerrak and the Kattegat. Details of the survey are given in the 'Report of the Planning Group for Herring Surveys' (ICES 2009/LRC:02). The 1999 was excluded due to different survey area coverage. The estimates of the Western Baltic spring spawning herring stock are 629000 tonnes and 8839 million individuals, which is similar to last year's estimate. The stock is dominated by 1 and 2 ringer fish. The results from this survey are summarised in Table 3.3.2.

### 3.3.3 Larvae Surveys

Herring larvae surveys in the western Baltic (Greifswalder Bodden and adjacent waters; SD 24) were conducted in weekly intervals during the 2008 spawning season (March to June). This was defined as the total number of larvae that reach the length of 20 mm (N20; Table 3.3.3) (Oeberst et al, 2007, WD 7 in HAWG 2008 (ICES 2008/ACOM:62)). The values estimated for 2008 is the lowest on record if we excluded the value observed in the first year of the time series (1992) and in line with the value observed in 2007 (Table 3.3.3).

### 3.4 Mean weights-at-age and maturity-at-age

Mean weights at age in the catch in the 1st quarter were used as stock weights (Table 3.2.14).

The maturity ogive of WBSS applied in HAWG has been assumed constant between years and thus been the same since 1991 (ICES 1992/Assess:13), although large year-to-year variations in the percentage mature have been observed (Gröhsler and Müller, 2004). A Workshop on Sexual Maturity Staging of Herring and Sprat is taking place during 2009 in order to, amongst other things, establish correspondence between old and new scales to convert time series and propose optimal sampling strategy to estimate accurate maturity ogives.

The same maturity ogive was used as in the HAWG 2008:

| W-RINGS | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8 +}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Maturity | 0.00 | 0.00 | 0.20 | 0.75 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 |

### 3.5 Recruitment

Indices of recruitment of 0-ringer western Baltic spring spawning herring (WBSS) in Subdivisions 22-24 for 2008 were available from the revised larval survey and are described in Section 3.3.3 and Oeberst et al., 2007 (WD 7 to the HAWG 2007(ICES 2007/ACFM:11)).

### 3.6 Assessment of Western Baltic spring spawners in Division IIIa and Subdivisions 22-24

### 3.6.1 Input data

### 3.6.1.1 Catch data

Catch in numbers at age from 1991 to 2008 were available for Subdivision IVa (East), Division IIIa and Subdivisions 22-24 (Table 3.6.1; Figure 3.6.1.1). Years before 1991 are excluded due to lack of reliable data for splitting spawning type and also due to a large change in fishing pattern caused by changes in the German fishing fleets (ICES 2008/ACOM:62).

Mean weights at age in the catch vary annually and are available for the same period as the catch in numbers (Table 3.6.2; Figure 3.6.1.3). Proportions at age (by weight) thus reflect the combined variation in numbers at age and weight at age (Figure 3.6.1.2).

### 3.6.1.2 Biological data

Estimates of the mean weight of individuals in the stock (Tables 3.2.14 and 3.6.3) are available for all years considered.

Natural mortality was assumed constant over time and equal to $0.3,0.5$, and 0.2 for $0-$ ringers, 1 -ringers, and $2+$-ringers respectively (Table 3.6.4). The estimates of natural mortality were derived as a mean for the years 1977-1995 from the Baltic MSVPA (ICES 1997/J:2).

The proportion of individuals that are mature is assumed constant over the period considered (Table 3.6.5): ages $0-1$ are assumed to be all immature, ages $2-4$ are $20 \%$, $75 \%$ and $90 \%$ mature respectively, and all older ages are $100 \%$ mature.

The proportions of fishing mortality, $\mathrm{F}(0.1)$ and natural mortality $\mathrm{M}(0.25)$ before spawning are assumed constant between years (Table 3.6.6-7). The difference between these two values arises due to the fact that the fishery is prosecuted in the latter half of the year.

### 3.6.1.3 Surveys

All surveys covering this stock were previously explored in terms of time series trends, internal consistency, and mortality signals during the Benchmark Assessment of this stock. The choice of age groups included was made there on the basis of existing knowledge of migration patterns and the analysis of the internal consistency of the surveys by age. (ICES 2008/ACOM:62; Payne et. al 2009) The final combination of surveys chosen was to include the N20 index as a recruitment index and apply the HERAS and German acoustic surveys to each characterise a subset of the total age classes.

The numerical values of the index for each individual age in each survey are given in Table 3.6.8, and are depicted in Figure 3.6.1.4. Each age and year in each survey is given an equal weighting.

### 3.6.2 Assessment method

As a part of the benchmark assessment process in 2008, the choice of assessment model was examined and the HAWG concluded that the underlying assumptions in the FLICA appeared to be valid. Details of the exact software package versions employed are given in Table 3.6.11.

### 3.6.3 Assessment configuration

Following the procedure in the WBSS stock annex (Annex 4), the following settings were used in this update assessment (Tables 3.6.9-10):

- The period for the separable constraint: 5 years (2003-2007)
- The weighing factor to all indices: lambda = 1
- A linear catchability model for all indices
- The reference $F$ set at age 4 and the selection=1 for the oldest age
- The catch data were down-weighted to 0.1 for 0 -ringer herring
- No stock-recruitment model was fitted
- Errors in index values are assumed to be correlated.
- Plus group is set to age $8+$.


### 3.6.4 Assessment Results

The results of the assessment are given in Tables 3.6.12-21. The estimated SSB for 2008 is 159406 tonnes. The mean fishing mortality (ages 3-6) is estimated as 0.37 . Parametric bootstrap estimates of these values give $95 \%$ confidence intervals of [112 000, 196 000 ] for SSB and [0.24, 0.55 ] for the mean fishing mortality (Figure 3.6.4.1).

After a marked decline from over 300000 tonnes in the early 1990s to a low of 120000 tonnes in the late 1990s, the SSB of this stock recovered somewhat, reaching a secondary peak of around 200000 tonnes in the early 2000s (Figure 3.6.4.2). After a small peak in 2006 coinciding with the maturing of the 2003 year-class the SSB has recently declined with about $17 \%$.

Fishing mortality on this stock was high in the mid 1990s, reaching a maximum of over $0.7 \mathrm{yr}^{-1}$. In recent years, the $\mathrm{F}_{3-6}$ value has stabilised around 0.4 (Figure 3.6.4.2).

Recruitment in 2008 is estimated at approximately 0.9 billion individuals. This is the lowest value observed in the 18 years covered by the assessment, and represents the continuation of a trend of decreasing recruitment from 2003 (Figure 3.6.4.2).

The catch residuals are generally free from patterns (Figure 3.6.4.3). The marginal totals of residuals between the catch and the separable model are small overall, although there does appear to be a trend in the age residuals on either side of the reference age.

The individual diagnostics for the three surveys generally show good quality fits (Figures 3.6.4.4-3.6.4.11). The residuals appear to be distributed randomly, and the assumption of their being distributed normally is generally held up. Systematic year effects appear to be present in the Herring acoustic survey (HERAS), especially in the later ages. Most survey-ages appear to have at least one significant outlier, often oc-
curring in the earlier part of the time series. Generally, however, the agreement between the data and the fitted model appears good through all data sources.

The mean contribution of the survey data points to the objective function is generally greater than that of the catch data (Figure 3.6.4.12): this is not surprising, however, as there are significantly more parameters fitted to the catch data. The agreement between the model and the GerAS survey is generally better than that of the HERAS survey. The N20 larval index shows the worst fit, on average.
Some patterns are apparent in the residuals (Figures 3.6.4.13). The HERAS survey shows appreciable year effects, with some years showing either positive or negative residuals across all ages. The German acoustic survey appears to give a more random pattern. The N20 index shows an improving fit in latter years, with one large dominating residual in its first year. The residuals are generally small (e.g. less than 0.5 ), but are dominated by a few outlying points. No cohort or age effects are apparent.

Retrospective analysis suggests the assessment method gives a relatively consistent perception of the stock and its development (Figure 3.6.4.14). There is a suggestion of biases in both the SSB and fishing mortality. However, the changes from year-to-year are generally less than the uncertainty of the estimated values (ICES 2008/ACOM:62), and are not a cause for great concern. The retrospective pattern in recruitment shows some variability, but is generally free from bias.

Retrospective analysis of the selectivity pattern for this fishery shows a stable selection pattern (Figure 3.6.4.15), especially in the most recent years covered by the separable period. Such a result suggests that the assumption of a constant selectivity in the fishery, a key criteria for the application of the FLICA method, is valid.

The stock-recruitment plot for this stock (Figure 3.6.4.16) does not show any clear relationship between stock-size and recruitment. Recent recruitment has dropped appreciably and consistently, while stock size has remained constant. The clustering of points suggests two different recruitment regimes, independent of stock size, may exist: a higher recruitment regime was present during the 1990s, which has been replaced by a lower recruitment regime in the 2000s. The mechanisms underpinning such a change remain unclear.

### 3.6.5 State of the stock

In the absence of defined reference points, the state of the stock cannot be evaluated. An analytical assessment demonstrates that the SBB has been stable over the last decade (to within $\pm 20 \%$ of the decadal mean), although the most recent value is in the lower quartile of all observations. Fishing mortality has also been stable in the same period but is larger than any proxy of $\mathrm{F}_{\mathrm{msy}}$. Recruitment has declined consistently since 2003 and the estimated number of 0-ringers in 2008 is the lowest observed value. These poor year classes have not had a dramatic effect on the spawning stock biomass as yet, due to the comparatively large size and good growth of the 2003 year class. This year class has been the largest component of the SSB for the last three years (2006-2008) and has supported the stock during this period (Figures 3.6.5.1-2). However, this year class is now in decline, and will pass out of the stock in the next two years, whilst its place will be taken by the sequence of poor year classes: a continuation of the decline in SSB can therefore be expected in the short and medium term.

### 3.6.6 Comparison with previous years perception of the stock

This year's assessment is an update assessment, and employs the same methodology as that in last year's Benchmark Assessment - the only difference between the two is the addition of a further year of data. The addition of this extra year of data appears to have modified the perception of the stock appreciably, increasing the SSBs and decreasing the fishing pressures estimated for 2006 and 2007 by around $20 \%$.

The text table below summarises the differences in the previous year's assessment configuration and perception of the stock.

| Category | Parameter | Assessment in 2009 | Assessment in 2008 | Diff. 08-09 <br> $(+/-) \%$ |
| :--- | :--- | :--- | :--- | :--- |
| ICA results | SSB 2006 <br> $\mathrm{F}(3-6) 2006$ | 192109 | 162978 | +17.9 |
|  | SSB 2007 <br> $\mathrm{F}(3-6) 2007$ | 0.396 | 0.491 | -19.3 |
|  | 161537 | 0.358 | 133503 | +21.0 |

### 3.7 Short term predictions

Short term predictions were made with the $f w d()$ method of "FLash" FLR package.

### 3.7.1 Input data

Stock numbers at age at the start of 2009 were taken from the ICA assessment, except for age 0 . For age 0 , the geometric mean recruitment (2002-2007) was assumed. The selection at age was taken from the ICA assessment. Arithmetic averages over the years 2006-2008 were used for mean weights at age in the catch and in the stock, as well as maturities at age. The input data are shown in Table 3.7.1.

### 3.7.2 Intermediate year 2009

A catch constraint was assumed for the intermediate year. The 2009 catch was estimated from the agreed TACs by fleet for Division IIIa and Subdivisions 22-24. The Division IIIa TAC includes both WBSS and NSAS herring, while the Subdivision 2224 TAC is assumed to be only WBSS herring.

- 2068 tonnes were subtracted from the Division IIIa TAC in 2008 and 1006 t subtracted from the TAC in 2009, to account for transfer of the Norwegian quota from IIIa to the North Sea.
- The catch by each fleet proportional to each TAC was assumed to be constant from 2008 to 2009. This gives the expected catch by fleet in 2009. Misreporting from Division IIIa into the North Sea is not explicitly included in these calculations, but is included implicitly via the proportions of quota taken.
- The catch of herring in Division IIIa consists of both WBSS and NSAS components. The expected catch of WBSS in IIIa was calculated assuming the same WBSS proportions in the catch of each fleet in 2009 as that in 2008 neglecting the small amount of about 120 t WBSS taken in Division IVaE by the A-fleet.

The resulting expected catch of WBSS in 2009 following this scheme was 45087 tonnes.

|  | 2008 |  |  |  |  | 2009 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calculation of Intermediate year (2009) catch constraint | Catch of WBSS | Catch of NSAS | $\begin{array}{r} \text { TAC* } \\ \text { (WBSS+ } \\ \text { NSAS) } \end{array}$ | Catch of NSAS <br> WBSS | Catch/ TAC | $\begin{array}{r} \text { TAC }^{*} \\ \text { in } \\ 2009 \end{array}$ | Total catch in 2009 | Proportion of WBSS in catch | Catch of WBSS in 2009 |
| A-fleet | 0.12 |  |  |  |  |  | 0.12 |  | 0.12 |
| C-fleet | 23.04 | 9.24 | 49.61 | 32.20 | 0.65 | 36.72 | 23.89 | 0.71 | 17.05 |
| D-fleet | 2.21 | 3.70 | 11.47 | 5.90 | 0.52 | 8.37 | 4.31 | 0.37 | 1.61 |
| F-fleet | 43.12 |  | 44.55 | 43.10 | 0.97 | 27.18 | 26.30 | 1.00 | 26.30 |
| Total (Div. IIIa, SD 22-24 and IVaE) | 68.48 |  |  |  |  |  | 54.63 |  | 45.087 |

*After accounting for Norwegian transfer from IIIa to North Sea (2 068 tonnes in 2008, 1006 tonnes in 2009).

### 3.7.3 Catch options for 2010

Detailed single option tables are presented for the following scenarios (Table 3.7.2).

## 1. Zero catch

After a decline in 2010 the SSB increases to 157000 t in 2011.
2. A $15 \%$ reduction of all fleet-wise TACs for 2009, converted into a total herring catch by assuming that the TAC is completely taken. The catches of WBSS herring are then calculated by assuming that the proportion of WBSS in each fleet's catch is the same as that in 2008, to give a catch in 2010 of 48100 t .

With this assumption the decline in SSB in 2010 continues in 2011 down to 114500 t, close to the suggested breakpoint of 110000 t .
3. As for option 2, but with no change in the TAC, to give a catch in 2010 of 56600 t .

With this assumption the decline in SSB in 2010 continues in 2011 down to 107400 t , below the suggested breakpoint of 110000 t .
4. As for option 2, but with a $15 \%$ increase in the TAC, to give a catch in 2010 of 65100 t .

With this assumption the decline in SSB in 2010 continues in 2011 down to 100300 t , well below the suggested breakpoint of 110000 t .
5. Catch in 2010 as assumed for 2009 (45 087 tonnes).

This option will lead to an SSB of 127700 t in 2010 and 117300 t 2011, a little above the breakpoint of 110000 tonnes.
6. $F_{2010}=0.25$, which is thought to lie close to $F_{\text {msy }}$.

This option will give a yield of 39800 t in 2010, with an SSB of 128200 t in 2010 and 121700 t in 2011.

### 3.8 Precautionary and yield based reference points

No precautionary reference points are defined for this stock. No new information was available (ICES 2008 ACOM:27).

### 3.9 Quality of the Assessment

There is retrospective bias with an underestimate of SSB and overestimate of F in the order of $20 \%$ (see Section 3.6.6).

The reason for this bias is not clear. Two factors are suggested, the relatively strong positive residuals across all ages in the 2008 HERAS acoustic survey, and the fact that a sequence of positive catch residuals in 2003 has now moved out of the separable period into the VPA region of the model. These changes acting together can be shown to cause the changes in the most recent years by the observed amount.

After the assessment was completed, an error was found in the input data: specifically, the total landings from Germany in Sub division 22 and 24 in 2008 were entered as 21800 tonnes, instead of 22800 tonnes (Table 3.1.1). Exploratory runs suggest that this error changes the SSB in 2008 by $0.1 \%$, and the mean fishing mortality in 2008 by $1.5 \%$. Such an error has an even smaller and therefore negligible effect on the catch advice in 2010. This error was discovered at a late stage. Given its small impact on the quality, accuracy and precision of the assessment and on the results of the projections, HAWG decided not to redo the assessment.

### 3.10 Management Considerations

## Quotas in Division Illa

The quota for the C-fleet and the by-catch quota for the D-fleet are set for both stocks of North Sea autumn spawners (NSAS) and Western Baltic spring spawners (WBSS) together (see Section 2.7).

## ICES catch predictions versus management TAC

ICES gives advice on catch options for the entire distribution of the two herring stocks separately, whereas herring is managed by areas (see the following text diagram).


## Development of a management plan for WBSS herring

ICES has explored management options under different assumptions of fishing mortality and recruitment using stochastic simulation with and without TAC constraints, including changes in selection pattern and different levels of uncertainty in the assessment. A proxy for $\mathrm{Fms} \mathrm{\gamma}_{\mathrm{m}}=0.25$, a SSB breakpoint of 110000 t equal to the lowest observed SSB below which the state of the stock is uncertain, and a maximum TAC variation of $+/-15 \%$ was suggested by WKHMP in 2008 (ICES 2008 ACOM:27). ICES concluded that, if recruitment does not further decline below the recent years' average, a fishing mortality of 0.25 could be a target in the development of a management plan for the western Baltic spring spawning herring stock.

## Data used for catch options in 2009 (intermediate year)

There is no firm basis for predicting the yearly fraction of NSAS in the catches of the C- and D-fleets. The proportions of the two stocks are influenced by the year class strength and their relative geographical distributions as well as fleet behaviour.

The procedure of deriving separate catches by stock and fleet is described in the stock annex for North Sea herring. The catch option for 2010 is based on the share by fleet and stock composition in catches for the most recent year 2008.

## Exploring a range of total WBSS catches for 2010 (prediction year)

Fleet wise catch options for the prediction year have the following assumptions:

- The TAC distribution by fleet in 2010 will be equal to 2009.
- There will be allowed a subtraction of $20 \%$ of the Norwegian quota that is transferred to the A-fleet (as NSAS).
- Each fleet catches its total TAC.
- The 2008 proportions of WBSS by fleet hold for 2010. (The proportions of WBSS in catches were 0.71 in the C-fleet, 0.37 in the D-fleet and 1.00 in the Ffleet).
- A constant catch of about 120 t of WBSS caught in the A-fleet in Division IVa East.

The table below gives the 2010 fleet wise catch options for the Western Baltic spring spawners and North Sea autumn spawners in Division IIIa, in Subdivisions 22-24, and in Subarea IVaE for the catch options described in section 3.7:

1) $\mathrm{F}=0$ not shown, 2) $\mathrm{F}_{-15 \% \mathrm{TAC}}=0.31$, 3) $\mathrm{F}_{\mathrm{TAC}}=0.37$, 4) $\mathrm{F}_{+15 \% \mathrm{TAC}}=0.44$, 5) $\mathrm{F}_{\text {catch } 09}=0.29$ and 6) $F_{M S Y}=0.25$.

Catch option for the WBSS and NSAS herring stock in 2010


* total catches of WBSS herring include a small constant catch of 120 t WBSS taken by the A-fleet in Div. IVa East
** total C-fleet catches in Division IIIa, the \% of the Norwegian quota that can be transferred to the North Sea is subtracted
${ }^{1)}$ Catches in 2010 relative to the TACs for 2009 in Division IIIa and Subdivisions 22-24; Fleet C: 37722 t, Fleet D: 8373 t, Fleet F: 27176 t.

Adopting a fishing mortality of 0.25 (proxy for $\mathrm{F}_{\text {msy }}$ ) as suggested by WKHMP (ICES 2008/ACOM:27) will to some degree reduce but not stop the present decline in SSB (Table 3.7.2). Catches corresponding to an F below 0.31 in 2010 should keep the SSB above the breakpoint of 110000 t in 2011.

The catches of WBSS in the C- and D-fleets comprise $37 \%$ of the total out-take of the WBSS stock, whereas the catches of NSAS in the same fleets only comprise $5 \%$ of the total out-take of the NSAS stock. Due to the state of the WBSS stock exhibiting a drastic decline in recruitment and negative development of the spawning stock biomass both stocks now need to be considered in the management. Thus the resulting catch options were also used as constraints for short term predictions for the NSAS herring (see Section 2.7).

### 3.11 Ecosystem considerations

Herring in Division IIIa and Subdivisions 22-24 are migratory. There are feeding migrations from the Western Baltic into more saline waters of Division IIIa and the eastern parts of Division IVa. There are indications from parasite infections that yet unknown proportions of sub-stocks spawning at the southern coast in the Baltic proper may perform similar migrations.

Similarly to the North Sea herring the Western Baltic herring has recently produced five poor year classes in a row. Indications suggest that the declining trend continues and that the 2008 year class is the lowest ever in the time series. In a recent recruitment analysis for different Baltic herring stocks, the Baltic Sea Index (BSI) reflecting Sea Surface Temperature (SST) was the main predictor for Western Baltic herring (Cardinale et al. 2009). There are no indications of systematic changes in growth or age at maturity, and a candidate key stage for reduced recruitment is probably the larval stage. Recruitment failure appears to have been initiated before the observed occurrence of the Ctenophore (Mnemiopsis leidyi) in the Western Baltic. The specific reasons for reduced larval survival are not known. Further investigation of the causes of the poor recruitment will require targeted research projects.

Table 3.1.1 WESTERN BALTIC HERRING.
Total landings (autumn \& spring spawners) in 1987-2008 in thousands of tonnes.
(Data provided by Working Group members 2008).

| Year | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Skagerrak |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 105.0 | 144.4 | 47.4 | 62.3 | 58.7 | 64.7 | 87.8 | 44.9 | 43.7 | 28.7 | 14.3 |
| Faroe Islands |  |  |  |  |  |  |  |  |  |  |  |
| Germany |  |  |  |  |  |  |  |  |  |  |  |
| Norway | 1.2 | 5.7 | 1.6 | 5.6 | 8.1 | 13.9 | 24.2 | 17.7 | 16.7 | 9.4 | 8.8 |
| Sweden | 51.2 | 57.2 | 47.9 | 56.5 | 54.7 | 88.0 | 56.4 | 66.4 | 48.5 | 32.7 | 32.9 |
| Total | 157.4 | 207.3 | 96.9 | 124.4 | 121.5 | 166.6 | 168.4 | 129.0 | 108.9 | 70.8 | 56.0 |
| Kattegat |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 46.6 | 76.2 | 57.1 | 32.2 | 29.7 | 33.5 | 28.7 | 23.6 | 16.9 | 17.2 | 8.8 |
| Sweden | 29.8 | 49.7 | 37.9 | 45.2 | 36.7 | 26.4 | 16.7 | 15.4 | 30.8 | 27.0 | 18.0 |
| Total | 76.4 | 125.9 | 95.0 | 77.4 | 66.4 | 59.9 | 45.4 | 39.0 | 47.7 | 44.2 | 26.8 |
| Sub. Div. 22+2 |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 32.5 | 33.1 | 21.7 | 13.6 | 25.2 | 26.9 | 38.0 | 39.5 | 36.8 | 34.4 | 30.5 |
| Germany | 53.1 | 54.7 | 56.4 | 45.5 | 15.8 | 15.6 | 11.1 | 11.4 | 13.4 | 7.3 | 12.8 |
| Poland | 8.0 | 6.6 | 8.5 | 9.7 | 5.6 | 15.5 | 11.8 | 6.3 | 7.3 | 6.0 | 6.9 |
| Sweden | 7.8 | 4.6 | 6.3 | 8.1 | 19.3 | 22.3 | 16.2 | 7.4 | 15.8 | 9.0 | 14.5 |
| Total | 101.4 | 99.0 | 92.9 | 76.9 | 65.9 | 80.3 | 77.1 | 64.6 | 73.3 | 56.7 | 64.7 |
| Sub. Div. 23 |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 0.8 | 0.1 | 1.5 | 1.1 | 1.7 | 2.9 | 3.3 | 1.5 | 0.9 | 0.7 | 2.2 |
| Sweden | 0.2 | 0.1 | 0.1 | 0.1 | 2.3 | 1.7 | 0.7 | 0.3 | 0.2 | 0.3 | 0.1 |
| Total | 1.0 | 0.2 | 1.6 | 1.2 | 4.0 | 4.6 | 4.0 | 1.8 | 1.1 | 1.0 | 2.3 |
| Grand Total | 336.2 | 432.4 | 286.4 | 279.9 | 257.8 | 311.4 | 294.9 | 234.4 | 231.0 | 172.7 | 149.8 |


| Y ear | $1998{ }^{(2)}$ | $1999{ }^{(2)}$ | 2000 | $2001{ }^{(5)}$ | $2002{ }^{(4)}$ | 2003 | 2004 | 2005 | $2006{ }^{(1,3)}$ | 2007 | $2008{ }^{(1)}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Skagerrak |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 10.3 | 10.1 | 16.0 | 16.2 | 26.0 | 15.5 | 11.8 | 14.8 | 5.2 | 3.6 | 3.9 |
| F aroe Islands |  |  |  |  |  |  |  | 0.4 |  |  | 0.0 |
| Germany |  |  |  |  |  | 0.7 | 0.5 | 0.8 | 0.6 | 0.5 | 1.6 |
| Norway | 8.0 | 7.4 | 9.7 |  |  |  |  |  |  | 3.5 | 4.0 |
| Sweden | 46.9 | 36.4 | 45.8 | 30.8 | 26.4 | 25.8 | 21.8 | 32.5 | 26.0 | 19.4 | 16.5 |
| Total | 65.2 | 53.9 | 71.5 | 47.0 | 52.3 | 42.0 | 34.1 | 48.5 | 31.8 | 26.9 | 26.0 |
| Kattegat |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 23.7 | 17.9 | 18.9 | 18.8 | 18.6 | 16.0 | 7.6 | 11.1 | 8.6 | 9.2 | 7.0 |
| Sweden | 29.9 | 14.6 | 17.3 | 16.2 | 7.2 | 10.2 | 9.6 | 10.0 | 10.8 | 11.2 | 5.2 |
| Total | 53.6 | 32.5 | 36.2 | 35.0 | 25.9 | 26.2 | 17.2 | 21.1 | 19.4 | 20.3 | 12.2 |
| Sub. Div. 22+24 |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 30.1 | 32.5 | 32.6 | 28.3 | 13.1 | 6.1 | 7.3 | 5.3 | 1.4 | 2.8 | 3.1 |
| Germany | 9.0 | 9.8 | 9.3 | 11.4 | 22.4 | 18.8 | 18.5 | 21.0 | 22.9 | 24.6 | $21.8^{(6)}$ |
| Poland | 6.5 | 5.3 | 6.6 | 9.3 | - | 4.4 | 5.5 | 6.3 | 5.5 | 2.9 | 5.5 |
| Sweden | 4.3 | 2.6 | 4.8 | 13.9 | 10.7 | 9.4 | 9.9 | 9.2 | 9.6 | 7.2 | 7.0 |
| Total | 49.9 | 50.2 | 53.3 | 62.9 | 46.2 | 38.7 | 41.2 | 41.8 | 39.4 | 37.6 | 37.5 |

Sub. Div. 23

| Denmark | 0.4 | 0.5 | 0.9 | 0.6 | 4.6 | 2.3 | 0.1 | 1.8 | 1.8 | 2.9 | 5.3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sweden | 0.3 | 0.1 | 0.1 | 0.2 |  | 0.2 | 0.3 | 0.4 | 0.7 |  | 0.3 |
| Total | 0.7 | 0.6 | 1.0 | 0.8 | 4.6 | 2.6 | 0.4 | 2.2 | 2.5 | 2.9 | 5.7 |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| Grand Total | 169.4 | 137.2 | 162.0 | 145.7 | 128.9 | 109.5 | 92.8 | 113.6 | 93.0 | 87.7 | 81.3 |

${ }^{(1)}$ Preliminary data.
${ }^{(2)}$ Revised data for 1998 and 1999
Bold= German revised data for 2001
${ }^{(3)} 2000$ tonnes of Danish landings are missing, see text section 3.1.2
${ }^{(4)}$ The Danish national management regime for herring and sprat fishery in Subdivision 22 was changed in 2002
${ }^{(5)}$ The total landings in Skagerrak have been updated for 1995-2001 due to Norwegian misreportings into Skagerrak.
${ }^{(6)}$ This value is incorrect but could not be corrected in subsequent calculations before the completion of the meeting. The correct value is 22.8

Table 3.1.2
WESTERN BALTIC HERRING.
Landings (SOP) in 2002-2008 by fleet and quarter (1000 t).

| Year | Quarter | Div. Illa |  | SD 22-24 | Div. IIIa +SD 22-24 |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | F leet C | Fleet D | Fleet F | Total |
| 2002 | 1 | 11.4 | 6.2 | 19.6 | 37.2 |
|  | 2 | 6.3 | 2.1 | 18.3 | 26.7 |
|  | 3 | 23.2 | 7 | 1.5 | 31.7 |
|  | 4 | 14.2 | 2.5 | 13.3 | 30.0 |
|  | Total | 55.1 | 17.8 | 52.7 | 125.6 |
| 2003 | 1 | 10.9 | 7 | 20.3 | 38.2 |
|  | 2 | 7.9 | 1.3 | 12.9 | 22.1 |
|  | 3 | 21.9 | 0.9 | 1.5 | 24.3 |
|  | 4 | 15 | 3.3 | 5.6 | 23.9 |
|  | Total | 55.7 | 12.5 | 40.3 | 108.5 |
| 2004 | 1 | 13.5 | 2.8 | 20.4 | 36.7 |
|  | 2 | 2.8 | 3.3 | 10.4 | 16.5 |
|  | 3 | 8.2 | 10.8 | 2.4 | 21.4 |
|  | 4 | 5.9 | 5.0 | 8.6 | 19.4 |
|  | Total | 30.3 | 22.0 | 41.7 | 93.9 |
| 2005 | 1 | 16.6 | 6.1 | 20.4 | 43.1 |
|  | 2 | 3.4 | 1.9 | 15.6 | 20.9 |
|  | 3 | 23.4 | 3.4 | 1.9 | 28.7 |
|  | 4 | 12.0 | 2.6 | 5.8 | 20.5 |
|  | Total | 55.4 | 14.1 | 43.7 | 113.3 |
| 2006 | 1 | 15.3 | 5.9 | 15.1 | 36.2 |
|  | 2 | 2.6 | 0.1 | 17.2 | 19.9 |
|  | 3 | 15.7 | 0.8 | 3.0 | 19.5 |
|  | 4 | 8.3 | 2.4 | 6.5 | 17.3 |
|  | Total | 41.9 | 9.3 | 41.9 | 93.0 |
| 2007 | 1 | 7.7 | 3.0 | 18.8 | 29.5 |
|  | 2 | 3.8 | 0.1 | 10.5 | 14.4 |
|  | 3 | 22.4 | 0.8 | 1.7 | 24.9 |
|  | 4 | 7.7 | 1.8 | 9.5 | 18.9 |
|  | Total | 41.6 | 5.7 | 40.5 | 87.7 |
| 2008 | 1 | 8.2 | 3.9 | 18.4 | 30.5 |
|  | 2 | 2.7 | 0.3 | 10.3 | 13.3 |
|  | 3 | 14.9 | 0.6 | 6.0 | 21.5 |
|  | 4 | 6.5 | 1.0 | 8.4 | 16.0 |
|  | Total | 32.3 | 5.9 | 43.1 | 81.3 |

Table 3.2.1 WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.
Division: Skagerrak Year: 2008 Country: All

| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 1.96 | 42 | 4.92 | 37 | 6.89 | 39 |
|  | 2 | 19.83 | 92 | 20.83 | 88 | 40.66 | 90 |
|  | 3 | 1.18 | 114 | 2.74 | 110 | 3.93 | 111 |
|  | 4 | 0.13 | 142 | 0.58 | 142 | 0.70 | 142 |
|  | 5 | 0.41 | 147 | 0.44 | 181 | 0.85 | 164 |
|  | 6 | 0.07 | 193 | 0.31 | 193 | 0.38 | 193 |
|  | 7 | 0.03 | 205 | 0.13 | 205 | 0.16 | 205 |
|  | 8+ | 0.03 | 193 | 0.13 | 193 | 0.16 | 193 |
|  | Total | 23.63 |  | 30.09 |  | 53.72 |  |
|  | SOP |  | 2,150 |  | 2,591 |  | 4,742 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 1.67 | 49 | 3.57 | 42 | 5.24 | 44 |
|  | 2 | 18.77 | 94 | 0.02 | 94 | 18.79 | 94 |
|  | 3 | 0.93 | 114 | 0.00 | 102 | 0.94 | 114 |
|  | 4 | 0.22 | 121 | 0.00 | 121 | 0.22 | 121 |
|  | 5 | 0.81 | 132 | 0.01 | 130 | 0.82 | 132 |
|  | 6 | 0.17 | 141 | 0.00 | 141 | 0.17 | 141 |
|  | 7 | 0.05 | 154 | 0.00 | 154 | 0.05 | 154 |
|  | 8+ | 0.07 | 164 | 0.00 | 164 | 0.07 | 164 |
|  | Total | 22.70 |  | 3.61 |  | 26.31 |  |
|  | SOP |  | 2,125 |  | 152 |  | 2,277 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.62 | 35 | 0.03 | 20 | 0.66 | 34 |
|  | 1 | 75.03 | 77 | 0.06 | 75 | 75.09 | 77 |
|  | 2 | 22.01 | 109 | 0.01 | 108 | 22.03 | 109 |
|  | 3 | 14.07 | 126 | 0.01 | 126 | 14.08 | 126 |
|  | 4 | 6.33 | 151 | 0.00 | 149 | 6.34 | 151 |
|  | 5 | 7.52 | 183 | 0.01 | 183 | 7.52 | 183 |
|  | 6 | 3.61 | 181 | 0.00 | 180 | 3.62 | 181 |
|  | 7 | 1.20 | 191 | 0.00 | 193 | 1.21 | 191 |
|  | 8+ | 0.56 | 198 | 0.00 | 194 | 0.56 | 198 |
|  | Total | 130.96 |  | 0.13 |  | 131.09 |  |
|  | SOP |  | 13,281 |  | 10 |  | 13,291 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 4.04 | 37 | 2.04 | 26 | 6.07 | 33 |
|  | 1 | 33.99 | 81 | 3.56 | 77 | 37.55 | 81 |
|  | 2 | 7.72 | 110 | 0.28 | 104 | 8.00 | 110 |
|  | 3 | 4.26 | 125 | 0.10 | 122 | 4.36 | 125 |
|  | 4 | 1.67 | 163 | 0.02 | 187 | 1.69 | 164 |
|  | 5 | 2.27 | 185 | 0.01 | 221 | 2.28 | 185 |
|  | 6 | 1.06 | 183 | 0.01 | 170 | 1.06 | 183 |
|  | 7 | 0.34 | 182 | 0.01 | 187 | 0.34 | 182 |
|  | 8+ | 0.15 | 213 |  |  | 0.15 | 213 |
|  | Total | 55.49 |  | 6.02 |  | 61.51 |  |
|  | SOP |  | 5,271 |  | 378 |  | 5,649 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 4.66 | 37 | 2.07 | 26 | 6.73 | 33 |
|  | 1 | 112.66 | 77 | 12.12 | 50 | 124.77 | 75 |
|  | 2 | 68.33 | 100 | 21.14 | 88 | 89.47 | 97 |
|  | 3 | 20.45 | 124 | 2.86 | 110 | 23.31 | 123 |
|  | 4 | 8.35 | 153 | 0.60 | 144 | 8.95 | 152 |
|  | 5 | 11.00 | 178 | 0.47 | 181 | 11.46 | 178 |
|  | 6 | 4.91 | 180 | 0.32 | 192 | 5.23 | 181 |
|  | 7 | 1.62 | 188 | 0.14 | 204 | 1.76 | 189 |
|  | 8+ | 0.81 | 198 | 0.13 | 193 | 0.94 | 197 |
|  | Total | 232.78 |  | 39.85 |  | 272.63 |  |
|  | SOP |  | 22,827 |  | 3,132 |  | 25,958 |

Table 3.2.2 WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.
Division: Kattegat Year: 2008 Country: ALL

| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 6.68 | 39 | 14.49 | 28 | 21.17 | 32 |
|  | 2 | 35.82 | 73 | 8.71 | 74 | 44.53 | 73 |
|  | 3 | 12.89 | 102 | 1.88 | 110 | 14.78 | 103 |
|  | 4 | 2.90 | 129 | 0.29 | 131 | 3.19 | 129 |
|  | 5 | 3.15 | 149 | 0.13 | 138 | 3.28 | 149 |
|  | 6 | 2.00 | 193 | 0.04 | 164 | 2.05 | 192 |
|  | 7 | 2.58 | 208 |  |  | 2.58 | 208 |
|  | 8+ | 0.28 | 201 | 0.01 | 158 | 0.29 | 199 |
|  | Total | 66.31 |  | 25.56 |  | 91.86 |  |
|  | SOP |  | 6,018 |  | 1,319 |  | 7,337 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 2.09 | 54 | 9.14 | 18 | 11.24 | 24 |
|  | 2 | 3.07 | 78 |  |  | 3.07 | 78 |
|  | 3 | 0.83 | 97 |  |  | 0.83 | 97 |
|  | 4 | 0.36 | 111 |  |  | 0.36 | 111 |
|  | 5 | 0.21 | 139 |  |  | 0.21 | 139 |
|  | 6 | 0.15 | 165 |  |  | 0.15 | 165 |
|  | 7 | 0.17 | 189 |  |  | 0.17 | 189 |
|  | 8+ | 0.05 | 164 |  |  | 0.05 | 164 |
|  | Total | 6.94 |  | 9.14 |  | 16.09 |  |
|  | SOP |  | 568 |  | 161 |  | 730 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $3$ | 0 | 0.20 | 22 | 42.44 | 15 | 42.64 | 15 |
|  | 1 | 16.94 | 63 | 0.00 | 65 | 16.94 | 63 |
|  | 2 | 3.47 | 79 | 0.00 | 84 | 3.47 | 79 |
|  | 3 | 1.08 | 105 | 0.00 | 115 | 1.08 | 105 |
|  | 4 | 0.59 | 105 | 0.00 | 137 | 0.59 | 105 |
|  | 5 | 0.20 | 180 | 0.00 | 180 | 0.20 | 180 |
|  | 6 | 0.15 | 129 | 0.00 | 204 | 0.15 | 129 |
|  | 7 | 0.07 | 211 | 0.00 | 140 | 0.07 | 211 |
|  | 8+ | 0.04 | 206 | 0.00 | 206 | 0.04 | 206 |
|  | Total | 22.74 |  | 42.44 |  | 65.18 |  |
|  | SOP |  | 1,599 |  | 631 |  | 2,230 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 0.30 | 36 | 40.90 | 15 | 41.20 | 15 |
|  | 1 | 13.66 | 67 | 0.58 | 73 | 14.24 | 67 |
|  | 2 | 2.36 | 82 | 0.09 | 92 | 2.45 | 82 |
|  | 3 | 0.68 | 106 | 0.03 | 122 | 0.72 | 107 |
|  | 4 | 0.46 | 106 | 0.01 | 145 | 0.48 | 107 |
|  | 5 | 0.04 | 147 | 0.00 | 147 | 0.04 | 147 |
|  | 6 | 0.12 | 108 | 0.00 | 120 | 0.12 | 108 |
|  | 7 | 0.06 | 236 | 0.00 | 238 | 0.06 | 236 |
|  | 8+ | 0.01 | 162 | 0.00 | 162 | 0.01 | 162 |
|  | Total | 17.70 |  | 41.62 |  | 59.32 |  |
|  | SOP |  | 1,271 |  | 666 |  | 1,938 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| T | 0 | 0.50 | 31 | 83.33 | 15 | 83.83 | 15 |
|  | 1 | 39.37 | 60 | 24.21 | 25 | 63.59 | 47 |
| 0 | 2 | 44.73 | 74 | 8.80 | 74 | 53.52 | 74 |
|  | 3 | 15.49 | 102 | 1.92 | 110 | 17.41 | 103 |
|  | 4 | 4.31 | 122 | 0.30 | 131 | 4.61 | 122 |
| t | 5 | 3.60 | 150 | 0.13 | 139 | 3.73 | 150 |
|  | 6 | 2.43 | 183 | 0.05 | 162 | 2.47 | 182 |
| a | 7 | 2.88 | 207 | 0.00 | 238 | 2.88 | 207 |
|  | 8+ | 0.37 | 195 | 0.02 | 158 | 0.39 | 194 |
|  | Total | 113.70 |  | 118.76 |  | 232.45 |  |
|  | SOP |  | 9,456 |  | 2,777 |  | 12,234 |

## Table 3.2.3 WESTERN BALTIC HERRING.

Landings in numbers (mill.), mean weight (g.) and SOP (t) by age and quarter.
Division: 22-24 Year: 2008 Country: ALL

| Quarter | W-rings | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 1.01 | 20 |  |  | 1.89 | 19 | 2.89 | 20 |
|  | 2 | 1.81 | 48 | 2.40 | 70 | 10.45 | 51 | 14.66 | 54 |
|  | 3 | 2.69 | 64 | 4.34 | 101 | 27.02 | 84 | 34.04 | 84 |
|  | 4 | 0.59 | 66 | 3.26 | 119 | 26.53 | 110 | 30.39 | 111 |
|  | 5 | 0.38 | 127 | 4.67 | 128 | 32.93 | 151 | 37.98 | 148 |
|  | 6 | 1.05 | 153 | 2.06 | 138 | 15.27 | 163 | 18.39 | 159 |
|  | 7 | 0.77 | 195 |  |  | 4.93 | 172 | 5.70 | 175 |
|  | 8+ | 1.97 | 205 | 0.90 | 164 | 6.73 | 185 | 9.59 | 187 |
|  | Total | 10.26 |  | 17.64 |  | 125.75 |  | 153.65 |  |
|  | SOP |  | 1,081 |  | 2,027 |  | 15,308 |  | 18,417 |
| Quarter | W-rings | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 2.66 | 13 | 0.38 | 25 | 3.51 | 26 | 6.55 | 21 |
|  | 2 | 0.14 | 48 | 1.05 | 75 | 19.38 | 48 | 20.57 | 49 |
|  | 3 | 0.21 | 64 | 0.88 | 90 | 30.48 | 71 | 31.57 | 71 |
|  | 4 | 0.05 | 66 | 0.62 | 120 | 24.13 | 87 | 24.80 | 88 |
|  | 5 | 0.03 | 127 | 0.39 | 119 | 17.55 | 100 | 17.97 | 101 |
|  | 6 | 0.08 | 153 | 0.31 | 152 | 12.45 | 117 | 12.84 | 119 |
|  | 7 | 0.06 | 195 | 0.08 | 163 | 4.40 | 131 | 4.55 | 132 |
|  | 8+ | 0.15 | 205 | 0.13 | 161 | 5.21 | 136 | 5.49 | 138 |
|  | Total | 3.37 |  | 3.84 |  | 117.13 |  | 124.34 |  |
|  | SOP |  | 115 |  | 370 |  | 9,796 |  | 10,281 |
| Quarter | W-rings | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
|  |  | Numbers Mean W. |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $3$ | 0 | 0.12 | 12 | 0.09 | 19 | 1.01 | 23 | 1.22 | 22 |
|  | 1 | 0.02 | 28 | 0.69 | 66 | 16.85 | 64 | 17.55 | 64 |
|  | 2 |  |  | 1.57 | 102 | 11.68 | 85 | 13.25 | 87 |
|  | 3 |  |  | 2.53 | 136 | 8.90 | 108 | 11.43 | 114 |
|  | 4 |  |  | 3.09 | 155 | 4.40 | 128 | 7.48 | 139 |
|  | 5 |  |  | 1.26 | 183 | 4.44 | 128 | 5.70 | 140 |
|  | 6 |  |  | 0.41 | 176 | 1.63 | 110 | 2.03 | 123 |
|  | 7 |  |  | 0.53 | 184 | 0.48 | 158 | 1.02 | 172 |
|  | 8+ |  |  | 0.54 | 224 | 0.13 | 215 | 0.67 | 222 |
|  | Total | 0.14 |  | 10.71 |  | 49.52 |  | 60.37 |  |
|  | SOP |  | 2 |  | 1,550 |  | 4,469 |  | 6,021 |
| Quarter |  | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 2.31 | 12 | 0.09 | 19 | 1.09 | 20 | 3.49 | 14 |
|  | 1 | 0.30 | 28 | 1.00 | 60 | 20.13 | 50 | 21.42 | 50 |
|  | 2 |  |  | 1.70 | 102 | 16.93 | 74 | 18.64 | 77 |
|  | 3 |  |  | 2.61 | 139 | 16.51 | 102 | 19.13 | 107 |
|  | 4 |  |  | 3.41 | 157 | 9.14 | 130 | 12.55 | 137 |
|  | 5 |  |  | 1.37 | 186 | 7.45 | 124 | 8.82 | 133 |
|  | 6 |  |  | 0.41 | 198 | 3.12 | 121 | 3.53 | 130 |
|  | 7 |  |  | 0.58 | 184 | 0.85 | 126 | 1.43 | 150 |
|  | 8+ |  |  | 0.62 | 220 | 0.66 | 140 | 1.28 | 178 |
|  | Total | 2.60 |  | 11.79 |  | 75.90 |  | 90.29 |  |
|  | SOP |  | 36 |  | 1,712 |  | 6,649 |  | 8,397 |
| Quarter | W-rings | Sub-division 22 |  | Sub-division 23 |  | Sub-division 24 |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| T | 0 | 2.43 | 12 | 0.17 | 19 | 2.11 | 21 | 4.71 | 16 |
|  | 1 | 3.98 | 16 | 2.07 | 56 | 42.38 | 52 | 48.42 | 50 |
|  | 2 | 1.95 | 48 | 6.72 | 86 | 58.44 | 63 | 67.12 | 65 |
| 0 | 3 | 2.89 | 64 | 10.36 | 118 | 82.92 | 85 | 96.17 | 88 |
|  | 4 | 0.64 | 66 | 10.38 | 142 | 64.20 | 106 | 75.22 | 111 |
| $t$ | 5 | 0.40 | 127 | 7.69 | 147 | 62.37 | 132 | 70.46 | 133 |
|  | 6 | 1.13 | 153 | 3.19 | 152 | 32.48 | 139 | 36.80 | 140 |
| a | 7 | 0.83 | 195 | 1.20 | 182 | 10.67 | 151 | 12.70 | 157 |
|  | 8+ | 2.12 | 205 | 2.19 | 194 | 12.73 | 163 | 17.04 | 172 |
|  | Total | 16.37 |  | 43.98 |  | 368.29 |  | 428.64 |  |
|  | SOP |  | 1,234 |  | 5,660 |  | 36,222 |  | 43,116 |

Table 3.2.4 HERRING IN DIVISION IIIa AND SUBDIVISIONS 22-24. Samples of commercial landings by quarter and area for 2008 available to the Working Group.

|  | Country | Quarter | $\begin{array}{r} \text { Landings } \\ \text { in '000 tons } \end{array}$ | Numbers of samples | Numbers of fish meas. | Numbers of fish aged |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Skagerrak | Denmark | 1 | 1.2 | 2 | 225 | 166 |
|  |  | 2 | 0.1 | 5 | 38 | - |
|  |  | 3 | 1.9 | 6 | 632 | 614 |
|  |  | 4 | 0.6 | 1 | 53 | 53 |
|  | Total |  | 3.9 | 14 | 948 | 833 |
|  | Germany | 1 | - | No data available |  |  |
|  |  | 2 | 0.0 |  |  |  |
|  |  | 3 | 0.2 |  |  |  |
|  |  | 4 | 1.4 |  |  |  |
|  | Total |  | 1.6 | 0 | 0 | 0 |
|  | Norway | 1 | 0.4 |  |  |  |
|  |  | 2 | 1.8 | No data available |  |  |
|  |  | 3 | 0.5 |  |  |  |
|  |  | 4 | 1.4 |  |  |  |
|  | Total |  | 4.0 | 0 | 0 | 0 |
|  | Sweden | 1 | 3.1 | 7 | 668 | 668 |
|  |  | 2 | 0.3 | 5 | 667 | 667 |
|  |  | 3 | 10.8 | 20 | 1,218 | 1,218 |
|  |  | 4 | 2.2 | 9 | 730 | 730 |
|  | Total |  | 16.5 | 41 | 3,283 | 3,283 |
| Kattegat | Denmark | 1 | 4.2 | 7 | 801 | 570 |
|  |  | 2 | 0.3 | 2 | 12 | 11 |
|  |  | 3 | 1.3 | 2 | 246 | 103 |
|  |  | 4 | 1.2 | No data available |  |  |
|  | Total |  | 7.0 | 11 | 1,059 | 684 |
|  | Sweden | 1 | 3.1 | 13 | 659 | 659 |
|  |  | 2 | 0.4 | 7 | 664 | 664 |
|  |  | 3 | 0.9 | 5 | 662 | 662 |
|  |  | 4 | 0.8 | 6 | 714 | 714 |
|  | Total |  | 5.2 | 31 | 2,699 | 2,699 |
| Subdivision 22 | Denmark | 1 | 0.0 | 4 | 169 | 169 |
|  |  | 2 | 0.0 | 2 | 166 | 63 |
|  |  | 3 | 0.0 | 2 | 175 | 77 |
|  |  | 4 | 0.0 | No data available |  |  |
|  | Total |  | 0.0 | 8 | 510 | 309 |
|  | Germany | 1 | 1.1 | No data available |  |  |
|  |  | 2 | 0.1 | 2 | 857 | 176 |
|  |  | 3 | 0.0 | No data available No data available |  |  |
|  |  | 4 | 0.0 |  |  |  |
|  | Total |  | 1.2 | 2 | 857 | 176 |
| Subdivision 23 | Denmark | 1 | 2.0 | 1 | 156 | 54 |
|  |  | 2 | 0.4 | 2 | 372 | 52 |
|  |  | 3 | 1.4 | No data available |  |  |
|  |  | 4 | 1.5 | 2 | 128 | 51 |
|  | Total |  | 5.3 | 5 | 656 | 157 |
|  | Sweden | 1 | 0.0 | No data available |  |  |
|  |  | 2 | 0.0 |  |  |  |
|  |  | 3 | 0.2 |  |  |  |
|  |  | 4 | 0.2 |  |  |  |
|  | Total |  | 0.3 | 0 | 0 | 0 |
| Subdivision 24 | Denmark | 1 | 1.8 | 6 | 827 | 394 |
|  |  | 2 | 0.5 |  | data available |  |
|  |  | 3 | 0.6 |  | data available |  |
|  |  | 4 | 0.2 | 3 | 407 | 214 |
|  | Total |  | 3.0 | 9 | 1,234 | 608 |
|  | Germany |  | 11.3 | 29 | 10,765 | 2,303 |
|  |  | 2 | 5.3 | 13 | 5,732 | 1,048 |
|  |  | 3 | 2 | No data available |  |  |
|  |  | 4 | 2.4 | 13 | 4,871 | 1,226 |
|  | Total |  | 20.6 | 55 | 21,368 | 4,577 |
|  | Poland | 1 | 0.6 | 4 | 395 | 169 |
|  |  | 2 | 2.9 | 7 | 3,141 | 408 |
|  |  | 3 | 1.4 | No data available No data available |  |  |
|  |  | 4 | 0.6 |  |  |  |
|  | Total |  | 5.5 | 11 | 3536 | 577 |
|  | Sweden | 1 | 1.6 | 5 | 649 | 649 |
|  |  | 2 | 1.1 | 2 | 491 | 491 |
|  |  | 3 | 0.8 | 1 | 256 | 256 |
|  |  | 4 | 3.5 | 6 | 707 | 707 |
|  | Total |  | 7.0 | 14 | 2,103 | 2,103 |

Table 3.2.5 HERRING IN DIVISION IIIa AND SUBDIVISIONS 22-24.
Samples of landings by quarter and area used to to estimate catch in numbers and mean weight by age for 2008.

|  | Country | Quarter | Fleet | Sampling |
| :---: | :---: | :---: | :---: | :---: |
| Skagerrak | Denmark | 1 | C | Danish sampling in Q1 |
|  |  | 2 | C | Danish sampling in Q1 |
|  |  | 3 | C | Danish sampling in Q3 |
|  |  | 4 | C | Danish sampling in Q3 |
|  | Germany | 1 | C | No landings |
|  |  | 2 | C | Danish sampling in Q1 |
|  |  | 3 | C | Danish sampling in Q3 |
|  |  | 4 | C | Danish sampling in Q3 |
|  | Sweden | 1 | C | Swedish sampling in Q1 |
|  |  | 2 | C | Swedish sampling in Q2 |
|  |  | 3 | C | Swedish sampling in Q3 |
|  |  | 4 | C | Swedish sampling in Q4 |
|  | Denmark | 1 | D | Swedish sampling in Q1 |
|  |  | 2 | D | Danish sampling in Q2 |
|  |  | 3 | D | Danish sampling in Q4 |
|  |  | 4 | D | Danish sampling in Q4 |
|  | Sweden | 1 | D | Swedish sampling in Q1 |
|  |  | 2 | D | Swedish sampling in Q2 |
|  |  | 3 | D | Danish sampling in Q3 |
|  |  | 4 | D | Swedish sampling in Q4 |
|  | Norway | 1 | C | Danish sampling in Q1 |
|  |  | 2 | C | Danish sampling in Q1 |
|  |  | 3 | C | Danish sampling in Q3 |
|  |  | 4 | C | Danish sampling in Q3 |
| Kattegat | Denmark | 1 | C | Danish sampling in Q1 |
|  |  | 2 | C | Danish sampling in Q1 |
|  |  | 3 | C | Danish sampling in Q3 |
|  |  | 4 | C | Danish sampling in Q3 |
|  | Sweden | 1 | C | Swedish sampling in Q1 |
|  |  | 2 | C | Swedish sampling in Q2 |
|  |  | 3 | C | Swedish sampling in Q3 |
|  |  | 4 | C | Swedish sampling in Q4 |
|  | Denmark | 1 | D | Danish sampling in Q1 |
|  |  | 2 | D | Danish sampling in Q2 |
|  |  | 3 | D | Danish sampling in Q3 |
|  |  | 4 | D | Danish sampling in Q3 |
|  | Sweden | 1 | D | Danish sampling in Q1 |
|  |  | 2 | D | No landings |
|  |  | 3 | D | Danish sampling in Q3 |
|  |  | 4 | D | Danish sampling in Q4 |

Fleet $\mathrm{C}=$ Human consumption, Fleet $\mathrm{D}=$ Industrial landings.

Table 3.2.5 continued. HERRING IN DIVISION IIIa AND SUBDIVISIONS 22-24. Samples of landings by quarter and area used to to estimate catch in numbers and mean weight by age for 2008.

|  | Country | Quarter | Fleet | Sampling |
| :---: | :---: | :---: | :---: | :---: |
| Subdivision 22 | Denmark | 1 | F | Danish sampling in Q1 |
|  |  | 2 | F | Danish sampling in Q2 |
|  |  | 3 | F | Danish sampling in Q3 |
|  |  | 4 | F | Danish sampling in Q3 |
|  | Germany | 1 | F | Danish sampling in Q2 |
|  |  | 2 | F | Danish sampling in Q2 |
|  |  | 3 | F | Danish sampling in Q3 |
|  |  | 4 | F | Danish sampling in Q4 |
| Subdivision 23 | Denmark | 1 | F | Danish sampling in Q1 |
|  |  | 2 | F | Danish sampling in Q2 |
|  |  | 3 | F | Danish sampling in Q4 |
|  |  | 4 | F | Danish sampling in Q4 |
|  | Sweden | 1 | F |  |
|  |  | 2 | F | Swedish sampling in Q2 in Sub-division 24 |
|  |  | 3 | F | Swedish sampling in Q3 in Sub-division 24 |
|  |  | 4 | F | Swedish sampling in Q4 in Sub-division 24 |
| Subdivision 24 | Denmark | 1 | F | Danish sampling in Q1 |
|  |  | 2 | F | Danish sampling in Q2 |
|  |  | 3 | F | Danish sampling in Q4 |
|  |  | 4 | F | Danish sampling in Q4 |
|  | Germany | 1 | F | German sampling in Q1 |
|  |  | 2 | F | German sampling in Q2 |
|  |  | 3 | F | German sampling in Q4 |
|  |  | 4 | F | German sampling in Q4 |
|  | Poland | 1 | F | Polish sampling in Q1 |
|  |  | 2 | F | Polish sampling in Q2 |
|  |  | 3 | F | Danish sampling in Q4 |
|  |  | 4 | F | Danish sampling in Q4 |
|  | Sweden | 1 | F | Swedish sampling in Q1 |
|  |  | 2 | F | Swedish sampling in Q2 |
|  |  | 3 | F | Swedish sampling in Q3 |
|  |  | 4 | F | Swedish sampling in Q4 |

Fleet $\mathrm{C}=$ Human consumption, Fleet $\mathrm{D}=$ Industrial landings, Fleet $\mathrm{F}=$ All landings from Subdiv.22-24.

Table 3.2.6 WESTERN BALTIC HERRING.
Proportion of North Sea autumn spawners and Baltic spring spawners given in \% in Skagerrak and Kattegat by age and quarter.
Year:
2008

| Quarter | W-rings | Skagerrak |  | Kattegat |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | North Sea autumn SP | Baltic Spring SP | North Sea autumn SP | Baltic Spring SP |
| 1 | 1 | 85.29\% | 14.71\% | 75.65\% | 24.35\% |
|  | 2 | 71.76\% | 28.24\% | 51.15\% | 48.85\% |
|  | 3 | 2.00\% | 98.00\% | 7.14\% | 92.86\% |
|  | 4 | 0.00\% | 100.00\% | 5.13\% | 94.87\% |
|  | 5 | 0.00\% | 100.00\% | 0.00\% | 100.00\% |
|  | 6 | 0.00\% | 100.00\% | 0.00\% | 100.00\% |
|  | 7 | 0.00\% | 100.00\% | 10.00\% | 90.00\% |
|  | 8 | 0.00\% | 100.00\% | 10.00\% | 90.00\% |
| Quarter | W-rings | Skagerrak |  | Kattegat |  |
|  |  | North Sea autumn SP | Baltic Spring SP | North Sea autumn SP | Baltic Spring SP |
| 2 | 1 | 93.88\% | 6.12\% | 88.46\% | 11.54\% |
|  | 2 | 65.31\% | 34.69\% | 69.39\% | 30.61\% |
|  | 3 | 24.00\% | 76.00\% | 10.00\% | 90.00\% |
|  | 4 | 6.25\% | 93.75\% | 0.00\% | 100.00\% |
|  | 5 | 5.88\% | 94.12\% | 0.00\% | 100.00\% |
|  | 6 | 5.88\% | 94.12\% | 0.00\% | 100.00\% |
|  | 7 | 5.88\% | 94.12\% | 0.00\% | 100.00\% |
|  | 8 | 5.88\% | 94.12\% | 0.00\% | 100.00\% |
| Quarter | W-rings | Skagerrak |  | Kattegat |  |
|  |  | North Sea autumn SP | Baltic Spring SP | North Sea autumn SP | Baltic Spring SP |
| 3 | 0 | 71.88\% | 28.13\% | 91.16\% | 8.84\% |
|  | 1 | 50.21\% | 49.79\% | 14.67\% | 85.33\% |
|  | 2 | 23.96\% | 76.04\% | 5.26\% | 94.74\% |
|  | 3 | 1.56\% | 98.44\% | 2.44\% | 97.56\% |
|  | 4 | 0.00\% | 100.00\% | 0.00\% | 100.00\% |
|  | 5 | 0.00\% | 100.00\% | 0.00\% | 100.00\% |
|  | 6 | 0.00\% | 100.00\% | 0.00\% | 100.00\% |
|  | 7 | 4.55\% | 95.45\% | 0.00\% | 100.00\% |
|  | 8 | 4.55\% | 95.45\% | 0.00\% | 100.00\% |
| Quarter | W-rings | Skagerrak |  | Kattegat |  |
|  |  | North Sea autumn SP | Baltic Spring SP | North Sea autumn SP | Baltic Spring SP |
| 4 | 0 | 84.38\% | 15.63\% | 100.00\% | $0.00 \%$$78.72 \%$ |
|  | 1 | 17.65\% | 82.35\% | 21.28\% |  |
|  | 2 | 2.33\% | 97.67\% | 0.00\% | 100.00\% |
|  | 3 | 4.35\% | 95.65\% | 0.00\% | 100.00\% |
|  | 4 | 4.35\% | 95.65\% | 0.00\% | 100.00\% |
|  | 5 | 4.35\% | 95.65\% | 0.00\% | 100.00\% |
|  | 6 | 4.35\% | 95.65\% | 0.00\% | 100.00\% |
|  | 7 | 4.35\% | 95.65\% | 0.00\% | 100.00\% |
|  | 8 | 4.35\% | 95.65\% | 0.00\% | 100.00\% |

Table 3.2.7 WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP (t) by age,
quarter and fleet.
North Sea Autumn spawners
Division: Kattegat Year:
2008 Country: All

| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 5.05 | 39 | 10.96 | 28 | 16.01 | 32 |
|  | 2 | 18.32 | 73 | 4.45 | 74 | 22.78 | 73 |
|  | 3 | 0.92 | 102 | 0.13 | 110 | 1.06 | 103 |
|  | 4 | 0.15 | 129 | 0.01 | 131 | 0.16 | 129 |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 | 0.26 | 208 |  |  | 0.26 | 208 |
|  | 8+ | 0.03 | 201 | 0.00 | 158 | 0.03 | 199 |
|  | Total | 24.73 |  | 15.57 |  | 40.30 |  |
|  | SOP |  | 1,711 |  | 652 |  | 2,364 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 1.85 | 54 | 8.09 | 18 | 9.94 | 24 |
|  | 2 | 2.13 | 78 |  |  | 2.13 | 78 |
|  | 3 | 0.08 | 97 |  |  | 0.08 | 97 |
|  | 4 |  |  |  |  |  |  |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | 8+ |  |  |  |  |  |  |
|  | Total | 4.07 |  | 8.09 |  | 12.16 |  |
|  | SOP |  | 275 |  | 143 |  | 418 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.18 | 22 | 38.69 | 15 | 38.87 | 15 |
|  | 1 | 2.48 | 63 | 0.00 | 65 | 2.48 | 63 |
|  | 2 | 0.18 | 79 | 0.00 | 84 | 0.18 | 79 |
|  | 3 | 0.03 | 105 | 0.00 | 115 | 0.03 | 105 |
|  | 4 |  |  |  |  |  |  |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | 8+ |  |  |  |  |  |  |
|  | Total | 2.88 |  | 38.69 |  | 41.56 |  |
|  | SOP |  | 178 |  | 575 |  | 753 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 0.30 | 36 | 40.90 | 15 | 41.20 | 15 |
|  | 1 | 2.91 | 67 | 0.12 | 73 | 3.03 | 67 |
|  | 2 |  |  |  |  |  |  |
|  | 3 |  |  |  |  |  |  |
|  | 4 |  |  |  |  |  |  |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | 8+ |  |  |  |  |  |  |
|  | Total | 3.21 |  | 41.02 |  | 44.23 |  |
|  | SOP |  | 205 |  | 618 |  | 822 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 0.48 | 31 | 79.58 | 15 | 80.06 | 15 |
|  | 1 | 12.30 | 53 | 19.17 | 24 | 31.47 | 35 |
|  | 2 | 20.64 | 74 | 4.45 | 74 | 25.09 | 74 |
|  | 3 | 1.03 | 102 | 0.13 | 110 | 1.17 | 102 |
|  | 4 | 0.15 | 129 | 0.01 | 131 | 0.16 | 129 |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 | 0.26 | 208 |  |  | 0.26 | 208 |
|  | 8+ | 0.03 | 201 | 0.00 | 158 | 0.03 | 199 |
|  | Total | 34.88 |  | 103.36 |  | 138.24 |  |
|  | SOP |  | 2,368 |  | 1,988 |  | 4,356 |

Table 3.2.8 WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.

North Sea Autumn spawners
Division:

## Skagerrak Year:

2008 Country:
All

| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 1.67 | 42 | 4.20 | 37 | 5.87 | 39 |
|  | 2 | 14.23 | 92 | 14.95 | 88 | 29.18 | 90 |
|  | 3 | 0.02 | 114 | 0.05 | 110 | 0.08 | 111 |
|  | 4 |  |  |  |  |  |  |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |
|  | 8+ |  |  |  |  |  |  |
|  | Total | 15.93 |  | 19.20 |  | 35.13 |  |
|  | SOP |  | 1,387 |  | 1,477 |  | 2,864 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 1.57 | 49 | 3.35 | 42 | 4.92 | 44 |
|  | 2 | 12.26 | 94 | 0.01 | 94 | 12.27 | 94 |
|  | 3 | 0.22 | 114 | 0.00 | 102 | 0.22 | 114 |
|  | 4 | 0.01 | 121 | 0.00 | 121 | 0.01 | 121 |
|  | 5 | 0.05 | 132 | 0.00 | 130 | 0.05 | 132 |
|  | 6 | 0.01 | 141 | 0.00 | 141 | 0.01 | 141 |
|  | 7 | 0.00 | 154 | 0.00 | 154 | 0.00 | 154 |
|  | 8+ | 0.00 | 164 | 0.00 | 164 | 0.00 | 164 |
|  | Total | 14.13 |  | 3.37 |  | 17.50 |  |
|  | SOP |  | 1,261 |  | 141 |  | 1,402 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.45 | 35 | 0.02 | 20 | 0.47 | 34 |
|  | 1 | 37.67 | 77 | 0.03 | 75 | 37.70 | 77 |
|  | 2 | 5.27 | 109 | 0.00 | 108 | 5.28 | 109 |
|  | 3 | 0.22 | 126 | 0.00 | 126 | 0.22 | 126 |
|  | 4 |  |  |  |  |  |  |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 | 0.05 | 191 | 0.00 | 193 | 0.05 | 191 |
|  | 8+ | 0.03 | 198 | 0.00 | 194 | 0.03 | 198 |
|  | Total | 43.70 |  | 0.05 |  | 43.75 |  |
|  | SOP |  | 3,530 |  | 3 |  | 3,533 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 3.41 | 37 | 1.72 | 26 | 5.13 | 33 |
|  | 1 | 6.00 | 81 | 0.63 | 77 | 6.63 | 81 |
|  | 2 | 0.18 | 110 | 0.01 | 104 | 0.19 | 110 |
|  | 3 | 0.19 | 125 | 0.00 | 122 | 0.19 | 125 |
|  | 4 | 0.07 | 163 | 0.00 | 187 | 0.07 | 164 |
|  | 5 | 0.10 | 185 | 0.00 | 221 | 0.10 | 185 |
|  | 6 | 0.05 | 183 | 0.00 | 170 | 0.05 | 183 |
|  | 7 | 0.01 | 182 | 0.00 | 187 | 0.01 | 182 |
|  | 8+ | 0.01 | 213 |  |  | 0.01 | 213 |
|  | Total | 10.01 |  | 2.36 |  | 12.37 |  |
|  | SOP |  | 698 |  | 95 |  | 793 |
| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 3.85 | 37 | 1.74 | 26 | 5.60 | 33 |
|  | 1 | 46.92 | 75 | 8.21 | 42 | 55.13 | 70 |
|  | 2 | 31.94 | 96 | 14.97 | 88 | 46.91 | 93 |
|  | 3 | 0.65 | 121 | 0.06 | 111 | 0.71 | 120 |
|  | 4 | 0.09 | 157 | 0.00 | 176 | 0.09 | 157 |
|  | 5 | 0.15 | 168 | 0.00 | 187 | 0.15 | 168 |
|  | 6 | 0.06 | 175 | 0.00 | 161 | 0.06 | 175 |
|  | 7 | 0.07 | 187 | 0.00 | 184 | 0.07 | 187 |
|  | 8+ | 0.04 | 197 | 0.00 | 172 | 0.04 | 197 |
|  | Total | 83.76 |  | 24.99 |  | 108.75 |  |
|  | SOP |  | 6,877 |  | 1,716 |  | 8,592 |

Table 3.2.9 WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP (t) by age, quarter and fleet.
Division: Kattegat Year: Baltic Spring spawners 2008 Country: All

| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 1.63 | 39 | 3.53 | 28 | 5.15 | 32 |
|  | 2 | 17.50 | 73 | 4.25 | 74 | 21.76 | 73 |
|  | 3 | 11.97 | 102 | 1.75 | 110 | 13.72 | 103 |
|  | 4 | 2.75 | 129 | 0.27 | 131 | 3.02 | 129 |
|  | 5 | 3.15 | 149 | 0.13 | 138 | 3.28 | 149 |
|  | 6 | 2.00 | 193 | 0.04 | 164 | 2.05 | 192 |
|  | 7 | 2.33 | 208 |  |  | 2.33 | 208 |
|  | 8+ | 0.25 | 201 | 0.01 | 158 | 0.26 | 199 |
|  | Total | 41.58 |  | 9.99 |  | 51.57 |  |
|  | SOP |  | 4,307 |  | 666 |  | 4,973 |
| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 0.24 | 54 | 1.05 | 18 | 1.30 | 24 |
|  | 2 | 0.94 | 78 |  |  | 0.94 | 78 |
|  | 3 | 0.75 | 97 |  |  | 0.75 | 97 |
|  | 4 | 0.36 | 111 |  |  | 0.36 | 111 |
|  | 5 | 0.21 | 139 |  |  | 0.21 | 139 |
|  | 6 | 0.15 | 165 |  |  | 0.15 | 165 |
|  | 7 | 0.17 | 189 |  |  | 0.17 | 189 |
|  | 8+ | 0.05 | 164 |  |  | 0.05 | 164 |
|  | Total | 2.87 |  | 1.05 |  | 3.93 |  |
|  | SOP |  | 293 |  | 19 |  | 312 |
| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.02 | 22 | 3.75 | 15 | 3.77 | 15 |
|  | 1 | 14.45 | 63 | 0.00 | 65 | 14.45 | 63 |
|  | 2 | 3.29 | 79 | 0.00 | 84 | 3.29 | 79 |
|  | 3 | 1.05 | 105 | 0.00 | 115 | 1.05 | 105 |
|  | 4 | 0.59 | 105 | 0.00 | 137 | 0.59 | 105 |
|  | 5 | 0.20 | 180 | 0.00 | 180 | 0.20 | 180 |
|  | 6 | 0.15 | 129 | 0.00 | 204 | 0.15 | 129 |
|  | 7 | 0.07 | 211 | 0.00 | 140 | 0.07 | 211 |
|  | 8+ | 0.04 | 206 | 0.00 | 206 | 0.04 | 206 |
|  | Total | 19.86 |  | 3.75 |  | 23.62 |  |
|  | SOP |  | 1,421 |  | 56 |  | 1,477 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 |  |  |  |  |  |  |
|  | 1 | 10.76 | 67 | 0.46 | 73 | 11.21 | 67 |
|  | 2 | 2.36 | 82 | 0.09 | 92 | 2.45 | 82 |
|  | 3 | 0.68 | 106 | 0.03 | 122 | 0.72 | 107 |
|  | 4 | 0.46 | 106 | 0.01 | 145 | 0.48 | 107 |
|  | 5 | 0.04 | 147 | 0.00 | 147 | 0.04 | 147 |
|  | 6 | 0.12 | 108 | 0.00 | 120 | 0.12 | 108 |
|  | 7 | 0.06 | 236 | 0.00 | 238 | 0.06 | 236 |
|  | 8+ | 0.01 | 162 | 0.00 | 162 | 0.01 | 162 |
|  | Total | 14.50 |  | 0.60 |  | 15.10 |  |
|  | SOP |  | 1,067 |  | 49 |  | 1,115 |
|  |  | Fleet C |  | Fleet D |  | Total |  |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 0.02 | 22 | 3.75 | 15 | 3.77 | 15 |
|  | 1 | 27.08 | 63 | 5.04 | 30 | 32.12 | 58 |
|  | 2 | 24.09 | 75 | 4.34 | 74 | 28.43 | 75 |
|  | 3 | 14.46 | 102 | 1.78 | 110 | 16.25 | 103 |
|  | 4 | 4.16 | 121 | 0.29 | 131 | 4.45 | 122 |
|  | 5 | 3.60 | 150 | 0.13 | 139 | 3.73 | 150 |
|  | 6 | 2.43 | 183 | 0.05 | 162 | 2.47 | 182 |
|  | 7 | 2.63 | 207 | 0.00 | 238 | 2.63 | 207 |
|  | 8+ | 0.35 | 195 | 0.01 | 158 | 0.36 | 193 |
|  | Total | 78.81 |  | 15.40 |  | 94.21 |  |
|  | SOP |  | 7,088 |  | 789 |  | 7,877 |

Table 3.2.10 WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP (t) by age,
quarter and fleet.
Division: Skagerrak Year:

## Baltic Spring spawners

 2008 Country: All| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 0.29 | 42 | 0.72 | 37 | 1.01 | 39 |
|  | 2 | 5.60 | 92 | 5.88 | 88 | 11.48 | 90 |
|  | 3 | 1.16 | 114 | 2.69 | 110 | 3.85 | 111 |
|  | 4 | 0.13 | 142 | 0.58 | 142 | 0.70 | 142 |
|  | 5 | 0.41 | 147 | 0.44 | 181 | 0.85 | 164 |
|  | 6 | 0.07 | 193 | 0.31 | 193 | 0.38 | 193 |
|  | 7 | 0.03 | 205 | 0.13 | 205 | 0.16 | 205 |
|  | 8+ | 0.03 | 193 | 0.13 | 193 | 0.16 | 193 |
|  | Total | 7.70 |  | 10.89 |  | 18.59 |  |
|  | SOP |  | 764 |  | 1,114 |  | 1,878 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 0.10 | 49 | 0.22 | 42 | 0.32 | 44 |
|  | 2 | 6.51 | 94 | 0.01 | 94 | 6.52 | 94 |
|  | 3 | 0.71 | 114 | 0.00 | 102 | 0.71 | 114 |
|  | 4 | 0.21 | 121 | 0.00 | 121 | 0.21 | 121 |
|  | 5 | 0.76 | 132 | 0.01 | 130 | 0.77 | 132 |
|  | 6 | 0.16 | 141 | 0.00 | 141 | 0.16 | 141 |
|  | 7 | 0.05 | 154 | 0.00 | 154 | 0.05 | 154 |
|  | 8+ | 0.07 | 164 | 0.00 | 164 | 0.07 | 164 |
|  | Total | 8.57 |  | 0.24 |  | 8.81 |  |
|  | SOP |  | 864 |  | 11 |  | 875 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.18 | 35 | 0.01 | 20 | 0.18 | 34 |
|  | 1 | 37.36 | 77 | 0.03 | 75 | 37.39 | 77 |
|  | 2 | 16.74 | 109 | 0.01 | 108 | 16.75 | 109 |
|  | 3 | 13.85 | 126 | 0.01 | 126 | 13.86 | 126 |
|  | 4 | 6.33 | 151 | 0.00 | 149 | 6.34 | 151 |
|  | 5 | 7.52 | 183 | 0.01 | 183 | 7.52 | 183 |
|  | 6 | 3.61 | 181 | 0.00 | 180 | 3.62 | 181 |
|  | 7 | 1.15 | 191 | 0.00 | 193 | 1.15 | 191 |
|  | 8+ | 0.53 | 198 | 0.00 | 194 | 0.53 | 198 |
|  | Total | 87.27 |  | 0.07 |  | 87.34 |  |
|  | SOP |  | 9,751 |  | 7 |  | 9,757 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 0.63 | 37 | 0.32 | 26 | 0.95 | 33 |
|  | 1 | 27.99 | 81 | 2.93 | 77 | 30.93 | 81 |
|  | 2 | 7.54 | 110 | 0.27 | 104 | 7.81 | 110 |
|  | 3 | 4.07 | 125 | 0.10 | 122 | 4.17 | 125 |
|  | 4 | 1.60 | 163 | 0.02 | 187 | 1.61 | 164 |
|  | 5 | 2.17 | 185 | 0.01 | 221 | 2.18 | 185 |
|  | 6 | 1.01 | 183 | 0.01 | 170 | 1.02 | 183 |
|  | 7 | 0.32 | 182 | 0.01 | 187 | 0.33 | 182 |
|  | 8+ | 0.14 | 213 |  |  | 0.14 | 213 |
|  | Total | 45.48 |  | 3.66 |  | 49.14 |  |
|  | SOP |  | 4,572 |  | 283 |  | 4,856 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 0.81 | 36 | 0.33 | 26 | 1.13 | 33 |
|  | 1 | 65.74 | 79 | 3.90 | 68 | 69.65 | 78 |
|  | 2 | 36.39 | 104 | 6.17 | 89 | 42.56 | 102 |
|  | 3 | 19.79 | 124 | 2.80 | 110 | 22.59 | 123 |
|  | 4 | 8.26 | 152 | 0.60 | 144 | 8.86 | 152 |
|  | 5 | 10.85 | 178 | 0.47 | 181 | 11.32 | 178 |
|  | 6 | 4.85 | 180 | 0.32 | 192 | 5.17 | 181 |
|  | 7 | 1.55 | 188 | 0.14 | 204 | 1.69 | 189 |
|  | 8+ | 0.77 | 198 | 0.13 | 193 | 0.91 | 197 |
|  | Total | 149.02 |  | 14.86 |  | 163.88 |  |
|  | SOP |  | 15,950 |  | 1,416 |  | 17,366 |

Table 3.2.11 WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP (t) by age,
quarter and fleet.
Division: Illa

North Sea Autumn spawners
Year: 2008 Country: All

| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 6.73 | 40 | 15.16 | 31 | 21.89 | 34 |
|  | 2 | 32.55 | 82 | 19.41 | 85 | 51.95 | 83 |
|  | 3 | 0.94 | 102 | 0.19 | 110 | 1.13 | 103 |
|  | 4 | 0.15 | 129 | 0.01 | 131 | 0.16 | 129 |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 | 0.26 | 208 |  |  | 0.26 | 208 |
|  | 8+ | 0.03 | 201 | 0.00 | 158 | 0.03 | 199 |
|  | Total | 40.66 |  | 34.77 |  | 75.43 |  |
|  | SOP |  | 3,098 |  | 2,130 |  | 5,227 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 3.42 | 52 | 11.44 | 25 | 14.86 | 31 |
|  | 2 | 14.39 | 91 | 0.01 | 94 | 14.40 | 91 |
|  | 3 | 0.31 | 109 | 0.00 | 102 | 0.31 | 109 |
|  | 4 | 0.01 | 121 | 0.00 | 121 | 0.01 | 121 |
|  | 5 | 0.05 | 132 | 0.00 | 130 | 0.05 | 132 |
|  | 6 | 0.01 | 141 | 0.00 | 141 | 0.01 | 141 |
|  | 7 | 0.00 | 154 | 0.00 | 154 | 0.00 | 154 |
|  | 8+ | 0.00 | 164 | 0.00 | 164 | 0.00 | 164 |
|  | Total | 18.20 |  | 11.46 |  | 29.65 |  |
|  | SOP |  | 1,536 |  | 283 |  | 1,820 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.63 | 31 | 38.71 | 15 | 39.34 | 15 |
|  | 1 | 40.16 | 76 | 0.03 | 75 | 40.19 | 76 |
|  | 2 | 5.46 | 108 | 0.00 | 108 | 5.46 | 108 |
|  | 3 | 0.25 | 123 | 0.00 | 126 | 0.25 | 123 |
|  | 4 |  |  |  |  |  |  |
|  | 5 |  |  |  |  |  |  |
|  | 6 |  |  |  |  |  |  |
|  | 7 | 0.05 | 191 | 0.00 | 193 | 0.05 | 191 |
|  | 8+ | 0.03 | 198 | 0.00 | 194 | 0.03 | 198 |
|  | Total | 46.57 |  | 38.74 |  | 85.31 |  |
|  | SOP |  | 3,708 |  | 578 |  | 4,286 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 3.71 | 37 | 42.62 | 15 | 46.32 | 17 |
|  | 1 | 8.91 | 76 | 0.75 | 76 | 9.66 | 76 |
|  | 2 | 0.18 | 110 | 0.01 | 104 | 0.19 | 110 |
|  | 3 | 0.19 | 125 | 0.00 | 122 | 0.19 | 125 |
|  | 4 | 0.07 | 163 | 0.00 | 187 | 0.07 | 164 |
|  | 5 | 0.10 | 185 | 0.00 | 221 | 0.10 | 185 |
|  | 6 | 0.05 | 183 | 0.00 | 170 | 0.05 | 183 |
|  | 7 | 0.01 | 182 | 0.00 | 187 | 0.01 | 182 |
|  | 8+ | 0.01 | 213 |  |  | 0.01 | 213 |
|  | Total | 13.21 |  | 43.38 |  | 56.59 |  |
|  | SOP |  | 903 |  | 713 |  | 1,615 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 4.34 | 36 | 81.32 | 15 | 85.66 | 16 |
|  | 1 | 59.21 | 71 | 27.38 | 29 | 86.60 | 58 |
|  | 2 | 52.58 | 87 | 19.43 | 85 | 72.00 | 86 |
|  | 3 | 1.68 | 109 | 0.19 | 110 | 1.88 | 109 |
|  | 4 | 0.23 | 139 | 0.02 | 133 | 0.25 | 139 |
|  | 5 | 0.15 | 168 | 0.00 | 187 | 0.15 | 168 |
|  | 6 | 0.06 | 175 | 0.00 | 161 | 0.06 | 175 |
|  | 7 | 0.33 | 203 | 0.00 | 184 | 0.33 | 203 |
|  | 8+ | 0.06 | 199 | 0.00 | 159 | 0.07 | 198 |
|  | Total | 118.64 |  | 128.35 |  | 246.99 |  |
|  | SOP |  | 9,245 |  | 3,704 |  | 12,949 |

Table 3.2.12 WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP (t) by age,
quarter and fleet.
Division: Illa

Year:
Baltic Spring spawners 2008 Country: All

| Quarter | W-rings | Fleet C |  | Fleet D |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 1 | 1 | 1.91 | 40 | 4.25 | 30 | 6.17 | 33 |
|  | 2 | 23.10 | 78 | 10.14 | 82 | 33.24 | 79 |
|  | 3 | 13.13 | 103 | 4.44 | 110 | 17.57 | 105 |
|  | 4 | 2.88 | 130 | 0.85 | 138 | 3.72 | 132 |
|  | 5 | 3.56 | 149 | 0.57 | 171 | 4.13 | 152 |
|  | 6 | 2.07 | 193 | 0.35 | 189 | 2.42 | 192 |
|  | 7 | 2.35 | 207 | 0.13 | 205 | 2.49 | 207 |
|  | 8+ | 0.28 | 200 | 0.15 | 190 | 0.42 | 197 |
|  | Total | 49.28 |  | 20.88 |  | 70.16 |  |
|  | SOP |  | 5,071 |  | 1,780 |  | 6,851 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 0.34 | 53 | 1.27 | 22 | 1.62 | 28 |
|  | 2 | 7.45 | 92 | 0.01 | 94 | 7.46 | 92 |
|  | 3 | 1.46 | 105 | 0.00 | 102 | 1.46 | 105 |
|  | 4 | 0.56 | 114 | 0.00 | 121 | 0.57 | 114 |
|  | 5 | 0.98 | 134 | 0.01 | 130 | 0.98 | 134 |
|  | 6 | 0.31 | 153 | 0.00 | 141 | 0.31 | 153 |
|  | 7 | 0.22 | 181 | 0.00 | 154 | 0.22 | 181 |
|  | 8+ | 0.12 | 164 | 0.00 | 164 | 0.12 | 164 |
|  | Total | 11.45 |  | 1.29 |  | 12.74 |  |
|  | SOP |  | 1,157 |  | 30 |  | 1,187 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 3 | 0 | 0.19 | 34 | 3.76 | 15 | 3.95 | 16 |
|  | 1 | 51.81 | 73 | 0.03 | 75 | 51.84 | 73 |
|  | 2 | 20.03 | 104 | 0.01 | 108 | 20.04 | 104 |
|  | 3 | 14.91 | 124 | 0.01 | 126 | 14.91 | 124 |
|  | 4 | 6.92 | 147 | 0.00 | 149 | 6.93 | 147 |
|  | 5 | 7.72 | 183 | 0.01 | 183 | 7.72 | 183 |
|  | 6 | 3.77 | 179 | 0.00 | 180 | 3.77 | 179 |
|  | 7 | 1.22 | 192 | 0.00 | 192 | 1.22 | 192 |
|  | 8+ | 0.57 | 199 | 0.00 | 195 | 0.57 | 199 |
|  | Total | 107.13 |  | 3.82 |  | 110.95 |  |
|  | SOP |  | 11,172 |  | 63 |  | 11,234 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 0.63 | 37 | 0.32 | 26 | 0.95 | 33 |
|  | 1 | 38.75 | 77 | 3.39 | 77 | 42.14 | 77 |
|  | 2 | 9.91 | 104 | 0.36 | 101 | 10.26 | 104 |
|  | 3 | 4.76 | 122 | 0.13 | 122 | 4.89 | 122 |
|  | 4 | 2.06 | 150 | 0.03 | 168 | 2.09 | 151 |
|  | 5 | 2.20 | 184 | 0.02 | 206 | 2.22 | 184 |
|  | 6 | 1.13 | 175 | 0.01 | 157 | 1.14 | 175 |
|  | 7 | 0.38 | 190 | 0.01 | 194 | 0.38 | 190 |
|  | 8+ | 0.16 | 209 | 0.00 | 162 | 0.16 | 208 |
|  | Total | 59.98 |  | 4.26 |  | 64.24 |  |
|  | SOP |  | 5,639 |  | 332 |  | 5,971 |
| Quarter |  | Fleet C |  | Fleet D |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| Total | 0 | 0.82 | 36 | 4.08 | 16 | 4.90 | 19 |
|  | 1 | 92.82 | 74 | 8.95 | 46 | 101.76 | 72 |
|  | 2 | 60.48 | 92 | 10.51 | 83 | 71.00 | 91 |
|  | 3 | 34.25 | 115 | 4.58 | 110 | 38.84 | 114 |
|  | 4 | 12.42 | 142 | 0.89 | 140 | 13.31 | 142 |
|  | 5 | 14.45 | 171 | 0.60 | 172 | 15.05 | 171 |
|  | 6 | 7.28 | 181 | 0.37 | 188 | 7.65 | 181 |
|  | 7 | 4.18 | 200 | 0.14 | 204 | 4.32 | 200 |
|  | 8+ | 1.12 | 197 | 0.15 | 190 | 1.27 | 196 |
|  | Total | 227.84 |  | 30.26 |  | 258.10 |  |
|  | SOP |  | 23,038 |  | 2,205 |  | 25,243 |

Table 3.2.13
WESTERN BALTIC HERRING.
Total catch in numbers (mill) and mean weight (g), SOP (tonnes) of Western Baltic Spring spawners in Division IIIa and the North Sea in the years 1992-2008.

|  | W-rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | Numbers | 109.08 | 246.00 | 321.85 | 174.02 | 154.47 | 78.33 | 55.83 | 17.91 | 8.53 | 1,166.03 |
|  | Mean W. | 13.9 | 44.1 | 87.0 | 112.9 | 136.2 | 166.3 | 183.5 | 194.4 | 203.6 |  |
|  | SOP | 1,516 | 10,841 | 27,986 | 19,653 | 21,035 | 13,030 | 10,243 | 3,481 | 1,737 | 109,523 |
| 1993 | Numbers | 161.25 | 371.50 | 315.82 | 219.05 | 94.08 | 59.43 | 40.97 | 21.71 | 8.22 | 1,292.03 |
|  | Mean W. | 15.1 | 25.9 | 81.4 | 127.5 | 150.1 | 171.1 | 195.9 | 209.1 | 239.0 |  |
|  | SOP | 2,435 | 9,612 | 25,696 | 27,936 | 14,120 | 10,167 | 8,027 | 4,541 | 1,966 | 104,498 |
| 1994 | Numbers | 60.62 | 153.11 | 261.14 | 221.64 | 130.97 | 77.30 | 44.40 | 14.39 | 8.62 | 972.19 |
|  | Mean W. | 20.2 | 42.6 | 94.8 | 122.7 | 150.3 | 168.7 | 194.7 | 209.9 | 220.2 |  |
|  | SOP | 1,225 | 6,524 | 24,767 | 27,206 | 19,686 | 13,043 | 8,642 | 3,022 | 1,898 | 106,013 |
| 1995 | Numbers | 50.31 | 302.51 | 204.19 | 97.93 | 90.86 | 30.55 | 21.28 | 12.01 | 7.24 | 816.86 |
|  | Mean W. | 17.9 | 41.5 | 97.8 | 138.0 | 163.1 | 198.5 | 207.0 | 228.8 | 234.3 |  |
|  | SOP | 902 | 12,551 | 19,970 | 13,517 | 14,823 | 6,065 | 4,404 | 2,747 | 1,696 | 76,674 |
| 1996 | Numbers | 166.23 | 228.05 | 317.74 | 75.60 | 40.41 | 30.63 | 12.58 | 6.73 | 5.63 | 883.60 |
|  | Mean W. | 10.5 | 27.6 | 90.1 | 134.9 | 164.9 | 186.6 | 204.1 | 208.5 | 220.2 |  |
|  | SOP | 1,748 | 6,296 | 28,618 | 10,197 | 6,665 | 5,714 | 2,568 | 1,402 | 1,241 | 64,449 |
| 1997 | Numbers | 25.97 | 73.43 | 158.71 | 180.06 | 30.15 | 14.15 | 4.77 | 1.75 | 2.31 | 491.31 |
|  | Mean W. | 19.2 | 49.7 | 76.7 | 127.2 | 154.4 | 175.8 | 184.4 | 192.0 | 208.0 |  |
|  | SOP | 498 | 3,648 | 12,176 | 22,913 | 4,656 | 2,489 | 879 | 337 | 480 | 48,075 |
| 1998 | Numbers | 36.26 | 175.14 | 315.15 | 94.53 | 54.72 | 11.19 | 8.72 | 2.19 | 2.09 | 699.98 |
|  | Mean W. | 27.8 | 51.3 | 71.5 | 108.8 | 142.6 | 171.7 | 194.4 | 184.2 | 230.0 |  |
|  | SOP | 1,009 | 8,980 | 22,542 | 10,287 | 7,804 | 1,922 | 1,695 | 403 | 481 | 55,121 |
| 1999 | Numbers | 41.34 | 190.29 | 155.67 | 122.26 | 43.16 | 22.21 | 4.42 | 3.02 | 2.40 | 584.77 |
|  | Mean W. | 11.5 | 51.0 | 83.6 | 114.9 | 121.2 | 145.2 | 169.6 | 123.8 | 152.3 |  |
|  | SOP | 477 | 9,698 | 13,012 | 14,048 | 5,232 | 3,225 | 749 | 373 | 366 | 47,179 |
| 2000 | Numbers | 114.83 | 318.22 | 302.10 | 99.88 | 50.85 | 18.76 | 8.21 | 1.35 | 1.40 | 915.60 |
|  | Mean W. | 22.6 | 31.9 | 67.4 | 107.7 | 140.2 | 170.0 | 157.0 | 185.0 | 210.1 |  |
|  | SOP | 2,601 | 10,145 | 20,357 | 10,756 | 7,131 | 3,189 | 1,288 | 249 | 294 | 56,010 |
| 2001 | Numbers | 121.68 | 36.63 | 208.10 | 111.08 | 32.06 | 19.67 | 9.84 | 4.17 | 2.42 | 545.65 |
|  | Mean W. | 9.0 | 51.2 | 76.2 | 108.9 | 145.3 | 171.4 | 188.2 | 187.2 | 203.3 |  |
|  | SOP | 1,096 | 1,875 | 15,863 | 12,093 | 4,657 | 3,371 | 1,852 | 780 | 492 | 42,079 |
| 2002 | Numbers | 69.63 | 577.69 | 168.26 | 134.60 | 53.09 | 12.05 | 7.48 | 2.43 | 2.02 | 1,027.26 |
|  | Mean W. | 10.2 | 20.4 | 78.2 | 117.7 | 143.8 | 169.8 | 191.9 | 198.2 | 215.5 |  |
|  | SOP | 709 | 11,795 | 13,162 | 15,848 | 7,632 | 2,046 | 1,435 | 481 | 435 | 53,544 |
| 2003 | Numbers | 52.11 | 63.02 | 182.53 | 65.45 | 64.37 | 21.47 | 6.26 | 4.35 | 1.81 | 461.38 |
|  | Mean W. | 13.0 | 37.4 | 76.5 | 113.3 | 132.7 | 142.2 | 153.5 | 169.9 | 162.2 |  |
|  | SOP | 678 | 2,355 | 13,957 | 7,416 | 8,540 | 3,053 | 961 | 740 | 294 | 37,994 |
| 2004 | Numbers | 25.67 | 209.34 | 96.02 | 93.98 | 18.24 | 16.84 | 4.51 | 1.51 | 0.59 | 466.71 |
|  | Mean W. | 27.1 | 43.2 | 81.9 | 117.1 | 145.4 | 157.4 | 170.7 | 184.4 | 187.1 |  |
|  | SOP | 695 | 9,047 | 7,869 | 11,005 | 2,652 | 2,651 | 769 | 279 | 111 | 35,078 |
| 2005 | Numbers | 95.3 | 96.9 | 203.3 | 75.4 | 46.9 | 9.3 | 11.5 | 3.5 | 1.4 | 543.51 |
|  | Mean W. | 14.1 | 54.9 | 85.6 | 121.6 | 148.3 | 162.7 | 176.3 | 178.3 | 200.6 |  |
|  | SOP | 1,341 | 5,319 | 17,415 | 9,163 | 6,961 | 1,519 | 2,028 | 618 | 282 | 44,645 |
| 2006 | Numbers | 7.3 | 104.1 | 115.6 | 114.2 | 48.9 | 55.7 | 11.1 | 10.3 | 5.2 | 472.49 |
| corrected | Mean W. | 16.6 | 36.9 | 82.9 | 113.0 | 142.5 | 175.2 | 198.2 | 209.5 | 220.0 |  |
|  | SOP | 121 | 3,847 | 9,584 | 12,907 | 6,972 | 9,765 | 2,199 | 2,159 | 1,134 | 48,688 |
| 2007 | Numbers | 1.6 | 103.9 | 90.9 | 36.9 | 30.8 | 12.8 | 9.4 | 6.2 | 2.7 | 295.2 |
|  | Mean W. | 25.2 | 65.6 | 85.0 | 115.7 | 138.4 | 159.2 | 190.8 | 178.6 | 211.9 |  |
|  | SOP | 41 | 6,816 | 7,723 | 4,269 | 4,265 | 2,035 | 1,802 | 1,114 | 567 | 28,632 |
| 2008 | Numbers | 4.9 | 101.8 | 71.1 | 38.9 | 13.5 | 15.1 | 7.7 | 4.5 | 1.3 | 258.8 |
|  | Mean W. | 19.2 | 71.5 | 91.1 | 114.5 | 142.2 | 171.2 | 181.4 | 200.0 | 196.4 | 98.0 |
|  | SOP | 94 | 7,281 | 6,472 | 4,456 | 1,917 | 2,590 | 1,402 | 900 | 256 | 25,368 |

Data for 1995 to 2001 was revised in 2003.

Table 3.2.14
WESTERN BALTIC HERRING.
Landings in numbers (mill.), mean weight (g.) and SOP ( $t$ )
by age and quarter from. Western Baltic Spring Spawners
(values from the North Sea, see Table 2.2.1-2.2.5)

|  |  | Division IV |  | Division Illa |  | Sub-division 22-24 |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quarter | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $1$ | 1 | 0.00 | 24.00 | 6.17 | 32.79 | 2.89 | 19.59 | 9.06 | 29 |
|  | 2 | 0.00 | 56.03 | 33.24 | 79.05 | 14.66 | 53.94 | 47.89 | 71 |
|  | 3 | 0.00 | 109.83 | 17.57 | 104.66 | 34.04 | 84.45 | 51.61 | 91 |
|  | 4 | 0.00 | 244.00 | 3.72 | 131.61 | 30.39 | 110.58 | 34.11 | 113 |
|  | 5 | 0.00 | 262.00 | 4.13 | 151.98 | 37.98 | 147.62 | 42.11 | 148 |
|  | 6 | 0.00 | 325.00 | 2.42 | 192.25 | 18.39 | 159.38 | 20.81 | 163 |
|  | 7 | 0.00 | 267.00 | 2.49 | 207.34 | 5.70 | 175.27 | 8.19 | 185 |
|  | 8+ | 0.00 | 268.50 | 0.42 | 196.72 | 9.59 | 187.33 | 10.02 | 188 |
|  | Total | 0.00 |  | 70.16 |  | 153.65 |  | 223.81 |  |
|  | SOP |  | 0 |  | 6,851 |  | 18,417 |  | 25,267 |
| Quarter | W-rings | Division IV |  | Division Illa |  | Sub-division 22-24 |  | Total |  |
|  |  | Numbers | Mean W. | Numbers ${ }^{\text {M }}$ Mean W. |  | Numbers | Mean W. | Numbers | Mean W. |
| 2 | 1 | 0.00 | 67.00 | 1.62 | 28.35 | 6.55 | 20.71 | 8.17 | 22 |
|  | 2 | 0.06 | 144.00 | 7.46 | 91.69 | 20.57 | 49.18 | 28.09 | 61 |
|  | 3 | 0.06 | 163.00 | 1.46 | 105.05 | 31.57 | 71.39 | 33.08 | 73 |
|  | 4 | 0.12 | 166.00 | 0.57 | 114.33 | 24.80 | 88.21 | 25.48 | 89 |
|  | 5 | 0.05 | 163.00 | 0.98 | 133.79 | 17.97 | 100.69 | 19.00 | 103 |
|  | 6 | 0.06 | 189.00 | 0.31 | 152.56 | 12.84 | 118.55 | 13.21 | 120 |
|  | 7 | 0.13 | 191.00 | 0.22 | 181.12 | 4.55 | 132.19 | 4.90 | 136 |
|  | 8+ | 0.02 | 210.43 | 0.12 | 164.09 | 5.49 | 138.37 | 5.64 | 139 |
|  | Total | 0.50 |  | 12.74 |  | 124.34 |  | 137.58 |  |
|  | SOP |  | 87 |  | 1,187 |  | 10,281 |  | 11,555 |
| Quarter |  | Division IV |  | Division IIIa |  | Sub-division 22-24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $3$ | 0 | 0.00 | 0.00 | 3.95 | 15.80 | 1.22 | 21.54 | 5.17 | 17 |
|  | 1 | 0.00 | 71.72 | 51.84 | 72.98 | 17.55 | 64.29 | 69.39 | 71 |
|  | 2 | 0.01 | 128.31 | 20.04 | 104.17 | 13.25 | 86.61 | 33.30 | 97 |
|  | 3 | 0.02 | 142.52 | 14.91 | 124.21 | 11.43 | 114.22 | 26.37 | 120 |
|  | 4 | 0.05 | 159.68 | 6.93 | 147.06 | 7.48 | 139.04 | 14.46 | 143 |
|  | 5 | 0.02 | 179.86 | 7.72 | 182.64 | 5.70 | 139.92 | 13.44 | 165 |
|  | 6 | 0.02 | 199.76 | 3.77 | 178.56 | 2.03 | 122.97 | 5.83 | 159 |
|  | 7 | 0.05 | 219.58 | 1.22 | 191.79 | 1.02 | 171.78 | 2.29 | 184 |
|  | 8+ | 0.01 | 228.86 | 0.57 | 198.54 | 0.67 | 222.02 | 1.25 | 211 |
|  | Total | 0.19 |  | 110.95 |  | 60.37 |  | 171.51 |  |
|  | SOP |  | 35 |  | 11,234 |  | 6,021 |  | 17,290 |
| Quarter |  | Division IV |  | Division IIIa |  | Sub-division 22-24 |  | Total |  |
|  | W-rings | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| 4 | 0 | 0.00 | 15.02 | 0.95 | 33.28 | 3.49 | 14.46 | 4.44 | 18 |
|  | 1 | 0.00 | 56.03 | 42.14 | 77.11 | 21.42 | 50.27 | 63.57 | 68 |
|  | 2 | 0.00 | 109.83 | 10.26 | 103.60 | 18.64 | 76.57 | 28.90 | 86 |
|  | 3 | 0.01 | 244.00 | 4.89 | 122.07 | 19.13 | 106.75 | 24.02 | 110 |
|  | 4 | 0.00 | 262.00 | 2.09 | 150.67 | 12.55 | 137.43 | 14.65 | 139 |
|  | 5 | 0.00 | 325.00 | 2.22 | 184.07 | 8.82 | 133.16 | 11.04 | 143 |
|  | 6 | 0.00 | 267.00 | 1.14 | 174.81 | 3.53 | 130.12 | 4.67 | 141 |
|  | 7 | 0.00 | 268.00 | 0.38 | 190.31 | 1.43 | 149.67 | 1.81 | 158 |
|  | 8+ | 0.00 | 271.50 | 0.16 | 208.22 | 1.28 | 178.33 | 1.44 | 182 |
|  | Total | 0.01 |  | 64.24 |  | 90.29 |  | 154.54 |  |
|  | SOP |  | 3 |  | 5,971 |  | 8,397 |  | 14,371 |
| Quarter | W-rings | Division IV |  | Division IIIa |  | Sub-division 22-24 |  | Total |  |
|  |  | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. | Numbers | Mean W. |
| $\begin{gathered} \mathrm{T} \\ \mathbf{0} \\ \mathrm{t} \\ \mathrm{a} \\ \mathrm{I} \end{gathered}$ | 0 | 0.00 | 0.00 | 4.90 | 19.19 | 4.71 | 16.29 | 9.61 | 18 |
|  | 1 | 0.00 | 0.00 | 101.76 | 71.54 | 48.42 | 49.52 | 150.19 | 64 |
|  | 2 | 0.07 | 141.82 | 71.00 | 91.01 | 67.12 | 65.22 | 138.18 | 79 |
|  | 3 | 0.09 | 164.67 | 38.84 | 114.37 | 96.17 | 88.14 | 135.09 | 96 |
|  | 4 | 0.17 | 165.72 | 13.31 | 141.91 | 75.22 | 110.52 | 88.70 | 115 |
|  | 5 | 0.08 | 167.96 | 15.05 | 171.25 | 70.46 | 133.22 | 85.59 | 140 |
|  | 6 | 0.08 | 192.17 | 7.65 | 181.27 | 36.80 | 140.31 | 44.53 | 147 |
|  | 7 | 0.18 | 199.41 | 4.32 | 200.07 | 12.70 | 156.68 | 17.20 | 168 |
|  | 8+ | 0.04 | 215.85 | 1.27 | 195.89 | 17.04 | 172.24 | 18.34 | 174 |
|  | Total | 0.70 |  | 258.10 |  | 428.64 |  | 687.43 |  |
|  | SOP |  | 124 |  | 25,243 |  | 43,116 |  | 68,484 |

Table 3.2.15 WESTERN BALTIC HERRING
Total catch in numbers (mill) of Western Baltic Spring Spawners in Division Illa+North Sea and in Sub-Divisions 22-24 in the years 1992-2008

| Year Area |  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | Div. IV+Div. IIIa | 109.1 | 246.0 | 321.9 | 174.0 | 154.5 | 78.3 | 55.8 | 17.9 | 8.5 | 1056.9 |
|  | Sub-div. 22-24 | 36.0 | 210.7 | 280.8 | 190.8 | 179.5 | 104.9 | 84.0 | 34.8 | 14.0 | 1099.5 |
| 1993 | Div. IV+Div. IIIa | 161.3 | 371.5 | 315.8 | 219.0 | 94.1 | 59.4 | 41.0 | 21.7 | 8.2 | 1130.8 |
|  | Sub-div. 22-24 | 44.9 | 159.2 | 180.1 | 196.1 | 166.9 | 151.1 | 61.8 | 42.2 | 16.3 | 973.7 |
| 1994 | Div. IV+Div. IIIa | 60.6 | 153.1 | 261.1 | 221.6 | 131.0 | 77.3 | 44.4 | 14.4 | 8.6 | 911.6 |
|  | Sub-div. 22-24 | 202.6 | 96.3 | 103.8 | 161.0 | 136.1 | 90.8 | 74.0 | 35.1 | 24.5 | 721.6 |
| 1995 | Div. IV+Div. IIIa | 50.3 | 302.5 | 204.2 | 97.9 | 90.9 | 30.6 | 21.3 | 12.0 | 7.2 | 816.9 |
|  | Sub-div. 22-24 | 491.0 | 1,358.2 | 233.9 | 128.9 | 104.0 | 53.6 | 38.8 | 20.9 | 13.2 | 1951.5 |
| 1996 | Div. IV+Div. IIIa | 166.2 | 228.1 | 317.7 | 75.6 | 40.4 | 30.6 | 12.6 | 6.7 | 5.6 | 883.6 |
|  | Sub-div. 22-24 | 4.9 | 410.8 | 82.8 | 124.1 | 103.7 | 99.5 | 52.7 | 24.0 | 19.5 | 917.1 |
| 1997 | Div. IV+Div. IIIa | 26.0 | 73.4 | 158.7 | 180.1 | 30.2 | 14.2 | 4.8 | 1.8 | 2.3 | 491.3 |
|  | Sub-div. 22-24 | 350.8 | 595.2 | 130.6 | 96.9 | 45.1 | 29.0 | 35.1 | 19.5 | 21.8 | 973.2 |
| 1998 | Div. IV+Div. IIIa | 36.3 | 175.1 | 315.1 | 94.5 | 54.7 | 11.2 | 8.7 | 2.2 | 2.1 | 700.0 |
|  | Sub-div. 22-24 | 513.5 | 447.9 | 115.8 | 88.3 | 92.0 | 34.1 | 15.0 | 13.2 | 12.0 | 818.4 |
| 1999 | Div. IV+Div. IIIa | 41.3 | 190.3 | 155.7 | 122.3 | 43.2 | 22.2 | 4.4 | 3.0 | 2.4 | 584.8 |
|  | Sub-div. 22-24 | 528.3 | 425.8 | 178.7 | 123.9 | 47.1 | 33.7 | 11.1 | 6.5 | 3.7 | 830.5 |
| 2000 | Div. IV+Div. IIIa | 114.83 | 318.22 | 302.10 | 99.88 | 50.85 | 18.76 | 8.21 | 1.35 | 1.40 | 915.6 |
|  | Sub-div. 22-24 | 37.7 | 616.3 | 194.3 | 86.7 | 77.8 | 53.0 | 30.1 | 12.4 | 9.3 | 1079.9 |
| 2001 | Div. IV+Div. IIIa | 121.7 | 36.6 | 208.1 | 111.1 | 32.1 | 19.7 | 9.8 | 4.2 | 2.4 | 545.6 |
|  | Sub-div. 22-24 | 634.6 | 486.5 | 280.7 | 146.8 | 76.0 | 48.7 | 29.3 | 14.1 | 4.3 | 1721.0 |
| 2002 | Div. IV+Div. IIIa | 69.6 | 577.7 | 168.3 | 134.6 | 53.1 | 12.0 | 7.5 | 2.4 | 2.0 | 1027.3 |
|  | Sub-div. 22-24 | 80.6 | 81.4 | 113.6 | 186.7 | 119.2 | 45.1 | 31.1 | 11.4 | 6.3 | 675.4 |
| 2003 | Div. IV+Div. IIIa | 52.1 | 63.0 | 182.5 | 64.0 | 62.2 | 20.3 | 5.9 | 3.8 | 1.6 | 455.5 |
|  | Sub-div. 22-24 | 1.4 | 63.9 | 82.3 | 95.8 | 125.1 | 82.2 | 22.9 | 13.1 | 7.0 | 493.6 |
| 2004 | Div. IV+Div. IIIa | 25.7 | 209.3 | 96.0 | 94.0 | 18.2 | 16.8 | 4.5 | 1.5 | 0.6 | 466.7 |
|  | Sub-div. 22-24 | 217.9 | 248.4 | 101.8 | 70.8 | 75.0 | 74.4 | 44.5 | 13.4 | 10.4 | 856.5 |
| 2005 | Div. IV+Div. IIIa | 95.3 | 96.9 | 203.3 | 75.4 | 46.9 | 9.3 | 11.5 | 3.5 | 1.4 | 543.5 |
|  | Sub-div. 22-24 | 11.6 | 207.6 | 115.9 | 102.5 | 83.5 | 51.3 | 54.2 | 27.8 | 11.2 | 665.5 |
| 2006 c | Div. IV+Div. IIIa | 7.3 | 104.1 | 115.6 | 114.2 | 48.9 | 55.7 | 11.1 | 10.3 | 5.2 | 472.5 |
|  | Sub-div. 22-24 | 0.6 | 44.8 | 72.1 | 119.0 | 101.7 | 43.0 | 31.4 | 22.1 | 12.2 | 446.8 |
| 2007 | Div. IV+Div. IIIa | 1.6 | 103.9 | 90.9 | 36.9 | 30.8 | 12.8 | 9.4 | 6.2 | 2.7 | 295.2 |
|  | Sub-div. 22-24 | 9.1 | 68.2 | 93.9 | 107.0 | 96.1 | 52.2 | 20.8 | 15.0 | 12.1 | 474.3 |
| 2008 | Div. IV+Div. IIIa | 4.9 | 101.8 | 71.1 | 38.9 | 13.5 | 15.1 | 7.7 | 4.5 | 1.3 | 258.8 |
|  | Sub-div. 22-24 | 4.7 | 48.4 | 67.1 | 96.2 | 75.2 | 70.5 | 36.8 | 12.7 | 17.0 | 428.6 |

Data for 1995-2001 for the North Sea and Div. IIIa was revised in 2003.
c values have been corrected in 2007

Table 3.2.16
WESTERN BALTIC HERRING.
Mean weight ( $\mathbf{g}$ ) and SOP (tons) of Western Baltic Spring Spawners in Division Illa+North Sea and in Sub-Divisions 22-24 in the years 1992-2008


Data for 1995-2001 for the North Sea and Div. IIIa was revised in 2003.
c values have been corrected in 2007

Table 3.2.17 WESTERN BALTIC HERRING.
Transfers of North Sea autumn spawners from Div. Illa to the North Sea Numbers ('O00) and mean weight, SOP in (tonnes) 1992-2008.

|  | W-Rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | Number | 2,298.4 | 1,408.8 | 220.3 | 22.1 | 10.4 | 6.6 | 2.9 | 1.0 | 0.4 | 3,970.9 |
|  | Mean W. | 12.3 | 51.8 | 84.2 | 131.4 | 162.0 | 173.4 | 185.3 | 198.4 | 201.2 |  |
|  | SOP | 28,159 | 72,985 | 18,557 | 2,907 | 1,683 | 1,143 | 533 | 200 | 84 | 126,251 |
| 1993 | Number | 2,795.4 | 2,032.5 | 237.6 | 26.5 | 7.7 | 3.6 | 2.7 | 2.2 | 0.7 | 5,109.0 |
|  | Mean W. | 12.5 | 28.6 | 79.7 | 141.4 | 132.3 | 233.4 | 238.5 | 180.6 | 203.1 |  |
|  | SOP | 34,903 | 58,107 | 18,939 | 3,749 | 1,016 | 850 | 647 | 390 | 133 | 118,734 |
| 1994 | Number | 481.6 | 1,086.5 | 201.4 | 26.9 | 6.0 | 2.9 | 1.6 | 0.4 | 0.2 | 1,807.5 |
|  | Mean W. | 16.0 | 42.9 | 83.4 | 110.7 | 138.3 | 158.6 | 184.6 | 199.1 | 213.9 |  |
|  | SOP | 7,723 | 46,630 | 16,790 | 2,980 | 831 | 460 | 287 | 75 | 37 | 75,811 |
| 1995 | Number | 1,144.5 | 1,189.2 | 161.5 | 13.3 | 3.5 | 1.1 | 0.6 | 0.4 | 0.3 | 2,514.4 |
|  | Mean W. | 11.2 | 39.1 | 88.3 | 145.7 | 165.5 | 204.5 | 212.2 | 236.4 | 244.3 |  |
|  | SOP | 12,837 | 46,555 | 14,267 | 1,940 | 573 | 225 | 133 | 86 | 65 | 76,680 |
| 1996 | Number | 516.1 | 961.1 | 161.4 | 17.0 | 3.4 | 1.6 | 0.7 | 0.4 | 0.3 | 1,661.9 |
|  | Mean W. | 11.0 | 23.4 | 80.2 | 126.6 | 165.0 | 186.5 | 216.1 | 216.3 | 239.1 |  |
|  | SOP | 5,697 | 22,448 | 12,947 | 2,151 | 565 | 307 | 145 | 77 | 66 | 44,403 |
| 1997 | Number | 67.6 | 305.3 | 131.7 | 21.2 | 1.7 | 0.8 | 0.2 | 0.1 | 0.1 | 528.7 |
|  | Mean W. | 19.3 | 47.7 | 68.5 | 124.4 | 171.5 | 184.7 | 188.7 | 188.7 | 192.4 |  |
|  | SOP | 1,304 | 14,571 | 9,025 | 2,643 | 285 | 146 | 40 | 16 | 25 | 28,057 |
| 1998 | Number | 51.3 | 745.1 | 161.5 | 26.6 | 19.2 | 3.0 | 3.1 | 1.2 | 0.5 | 1,011.6 |
|  | Mean W. | 27.4 | 56.4 | 79.8 | 117.8 | 162.9 | 179.7 | 197.2 | 178.9 | 226.3 |  |
|  | SOP | 1,409 | 41,994 | 12,896 | 3,137 | 3,136 | 547 | 608 | 211 | 108 | 64,045 |
| 1999 | Number | 598.8 | 303.0 | 148.6 | 47.2 | 13.4 | 6.2 | 1.2 | 0.5 | 0.5 | 1,119.4 |
|  | Mean W. | 10.4 | 50.5 | 87.7 | 113.7 | 137.4 | 156.5 | 188.1 | 187.3 | 198.8 |  |
|  | SOP | 6,255 | 15,297 | 13,037 | 5,369 | 1,841 | 974 | 230 | 90 | 92 | 43,186 |
| 2000 | Number | 235.3 | 984.3 | 116.0 | 21.9 | 22.9 | 7.5 | 3.3 | 0.6 | 0.1 | 1,391.8 |
|  | Mean W. | 21.3 | 28.5 | 76.1 | 108.8 | 163.1 | 190.3 | 183.9 | 189.4 | 200.2 |  |
|  | SOP | 5,005 | 28,012 | 8,825 | 2,377 | 3,731 | 1,436 | 601 | 114 | 13 | 50,115 |
| 2001 | Number | 807.8 | 563.6 | 150.0 | 17.2 | 1.4 | 0.3 | 0.5 | 0.0 | 0.0 | 1,540.8 |
|  | Mean W. | 8.7 | 49.4 | 75.3 | 108.2 | 130.1 | 147.1 | 219.1 | 175.8 | 198.1 |  |
|  | SOP | 7,029 | 27,849 | 11,300 | 1,856 | 177 | 43 | 109 | 8 | 5 | 48,376 |
| 2002 | Number | 478.5 | 362.6 | 56.7 | 5.6 | 0.7 | 0.2 | 0.1 | 0.0 | 0.0 | 904.5 |
|  | Mean W. | 12.2 | 38.0 | 100.6 | 121.5 | 142.7 | 160.9 | 178.7 | 177.4 | 218.6 |  |
|  | SOP | 5,859 | 13,790 | 5,705 | 684 | 106 | 26 | 21 | 8 | 5 | 26,205 |
| 2003 | Number | 21.6 | 445.0 | 182.3 | 13.0 | 16.2 | 1.8 | 1.1 | 1.2 | 0.2 | 682.4 |
|  | Mean W. | 20.5 | 33.7 | 67.0 | 123.2 | 150.3 | 163.5 | 190.2 | 214.6 | 186.8 |  |
|  | SOP | 442 | 14,992 | 12,219 | 1,606 | 2,436 | 293 | 213 | 264 | 33 | 32,498 |
| 2004 | Number | 88.4 | 70.9 | 179.9 | 20.7 | 6.0 | 9.7 | 1.8 | 2.0 | 0.9 | 380.4 |
|  | Mean W. | 22.5 | 55.3 | 70.2 | 120.6 | 140.9 | 151.7 | 170.6 | 186.6 | 178.5 |  |
|  | SOP | 1,993 | 3,921 | 12,638 | 2,498 | 851 | 1,479 | 312 | 367 | 154 | 24,214 |
| 2005 | Number | 96.4 | 307.5 | 159.2 | 16.2 | 5.4 | 2.4 | 2.3 | 0.5 | 0.2 | 589.9 |
|  | Mean W. | 16.5 | 50.5 | 71.0 | 105.9 | 154.6 | 173.5 | 184.5 | 200.2 | 208.9 |  |
|  | SOP | 1,595 | 15,527 | 11,304 | 1,712 | 828 | 412 | 420 | 95 | 34 | 31,927 |
| 2006 | Number | 35.1 | 150.1 | 50.2 | 10.2 | 3.3 | 3.3 | 0.6 | 0.4 | 0.2 | 253.3 |
|  | Mean W. | 14.3 | 53.5 | 79.2 | 117.6 | 140.2 | 185.5 | 190.4 | 215.6 | 206.9 |  |
|  | SOP | 503 | 8,035 | 3,975 | 1,200 | 456 | 620 | 107 | 81 | 37 | 15,015 |
| 2007 | Number | 67.7 | 189.3 | 76.9 | 2.1 | 0.4 | 1.4 | 0.3 | 0.6 | 0.0 | 338.7 |
|  | Mean W. | 26.7 | 62.6 | 71.1 | 108.1 | 124.4 | 151.7 | 183.7 | 174.7 | 153.8 |  |
|  | SOP | 1,807 | 11,857 | 5,464 | 224 | 55 | 219 | 48 | 110 | 3 | 19,788 |
| 2008 | Number | 85.7 | 86.6 | 72.0 | 1.9 | 0.3 | 0.1 | 0.1 | 0.3 | 0.1 | 247.0 |
|  | Mean W. | 16.2 | 57.6 | 86.4 | 109.1 | 138.7 | 167.7 | 175.4 | 203.1 | 197.7 |  |
|  | SOP | 1,386 | 4,986 | 6,222 | 205 | 35 | 25 | 10 | 67 | 13 | 12,949 |

Corrections for the years 1991-1998 was made in WG2001, but are NOT included in the North Sea assessment.

Table 3.3.1 WESTERN BALTIC HERRING. German Acoustic survey (GerAS) on the Spring Spawning Herring in Sub-divisions 21 (Southern Kattegat, 41G0-42G2) - 24 in autumn 1993-2008 (September/October).

| Year | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001* 2 | 2002** | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Numbers in millions |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 893 | 5,475 | 5,108 | 1,833 | 2,859 | 2,490 | 5,994 | 1,009 | 2,478 | 4,103 | 3,777 | 2,555 | 3,055 | 4,159 | 2,591 | 2,150 |
| 1 | 492 | 416 | 1,675 | 1,439 | 1,955 | 801 | 1,339 | 1,430 | 1,126 | 838 | 1,238 | 969 | 753 | 950 | 560 | 393 |
| 2 | 437 | 884 | 329 | 590 | 738 | 679 | 287 | 454 | 1,227 | 421 | 223 | 592 | 640 | 274 | 278 | 214 |
| 3 | 530 | 560 | 358 | 434 | 395 | 394 | 233 | 329 | 845 | 575 | 217 | 346 | 401 | 376 | 149 | 209 |
| 4 | 403 | 444 | 354 | 295 | 162 | 237 | 156 | 202 | 367 | 341 | 260 | 163 | 192 | 353 | 136 | 150 |
| 5 | 125 | 189 | 254 | 306 | 119 | 100 | 52 | 79 | 132 | 64 | 97 | 143 | 105 | 183 | 88 | 166 |
| 6 | 55 | 60 | 127 | 119 | 99 | 51 | 8 | 39 | 86 | 25 | 38 | 79 | 90 | 131 | 25 | 102 |
| 7 | 28 | 24 | 46 | 47 | 33 | 24 | 1 | 6 | 20 | 10 | 9 | 23 | 26 | 85 | 23 | 42 |
| 8+ | 13 | 2 | 27 | 19 | 48 | 9 | 2 | 4 | 10 | 13 | 10 | 12 | 17 | 30 | 11 | 19 |
| Total | 2,976 | 8,053 | 8,277 | 5,083 | 6,409 | 4,785 | 8,072 | 3,551 | 6,290 | 6,389 | 5,869 | 4,882 | 5,279 | 6,542 | 3,860 | 3,445 |
| 3+ group | 1,154 | 1,279 | 1,166 | 1,220 | 856 | 815 | 452 | 658 | 1,459 | 1,028 | 631 | 766 | 830 | 1,159 | 432 | 688 |

Biomass ('000 tonnnes)

| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 12.8 | 66.9 | 58.5 | 16.6 | 28.5 | 23.8 | 71.8 | 13.8 | 31.2 | 38.2 | 33.9 | 23.1 | 33.1 | 43.9 | 25.8 | 24.8 |
| 1 | 19.5 | 14.5 | 58.6 | 46.6 | 76.4 | 39.9 | 51.1 | 57.5 | 48.2 | 34.2 | 44.8 | 35.9 | 30.1 | 38.8 | 23.0 | 17.7 |
| 2 | 21.7 | 41.0 | 20.9 | 29.1 | 43.5 | 50.1 | 22.0 | 28.4 | 75.9 | 30.0 | 16.1 | 34.5 | 48.6 | 19.7 | 20.8 | 12.5 |
| 3 | 33.8 | 40.7 | 30.1 | 31.0 | 35.9 | 35.3 | 27.5 | 27.7 | 77.2 | 56.8 | 22.0 | 27.7 | 36.2 | 35.9 | 12.6 | 17.7 |
| 4 | 25.7 | 43.0 | 40.1 | 21.2 | 22.3 | 28.0 | 16.7 | 24.1 | 38.0 | 40.4 | 34.2 | 18.4 | 22.7 | 37.4 | 12.5 | 14.3 |
| 5 | 12.7 | 24.2 | 27.3 | 37.1 | 16.7 | 11.4 | 6.8 | 9.3 | 18.5 | 9.0 | 14.6 | 17.3 | 14.4 | 27.2 | 8.9 | 16.8 |
| 6 | 7.1 | 12.3 | 14.9 | 16.1 | 14.0 | 6.2 | 0.9 | 5.6 | 13.3 | 3.5 | 5.7 | 12.2 | 14.5 | 19.9 | 2.9 | 8.8 |
| 7 | 2.3 | 5.3 | 9.3 | 6.1 | 5.3 | 3.7 | 0.3 | 1.2 | 3.9 | 1.1 | 1.3 | 3.4 | 5.2 | 14.6 | 2.6 | 3.5 |
| 8+ | 1.8 | 0.6 | 6.6 | 2.9 | 10.6 | 2.2 | 0.5 | 0.8 | 2.1 | 1.9 | 1.6 | 2.0 | 3.6 | 6.5 | 1.9 | 2.0 |
| Total | 137.3 | 248.5 | 266.3 | 206.8 | 253.3 | 200.5 | 197.5 | 168.4 | 308.1 | 215.0 | 174.2 | 174.6 | 208.3 | 243.9 | 111.0 | 118.0 |
| 3+ group | 83.3 | 126.2 | 128.2 | 114.4 | 104.9 | 86.8 | 52.6 | 68.7 | 152.9 | 112.6 | 79.4 | 81.1 | 96.5 | 141.5 | 41.4 | 63.0 |

Mean weight (g)

| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 14.3 | 12.2 | 11.5 | 9.0 | 10.0 | 9.5 | 12.0 | 13.7 | 12.6 | 9.3 | 9.0 | 9.0 | 10.8 | 10.5 | 10.0 | 11.5 |
| 1 | 39.7 | 34.8 | 35.0 | 32.4 | 39.1 | 49.8 | 38.2 | 40.2 | 42.8 | 40.8 | 36.2 | 37.0 | 40.0 | 40.8 | 41.0 | 45.0 |
| 2 | 49.7 | 46.4 | 63.7 | 49.4 | 58.9 | 73.8 | 76.6 | 62.6 | 61.8 | 71.1 | 72.3 | 58.3 | 76.0 | 71.9 | 74.8 | 58.4 |
| 3 | 63.9 | 72.8 | 84.1 | 71.5 | 91.1 | 89.5 | 118.2 | 84.3 | 91.4 | 98.7 | 101.3 | 80.1 | 90.2 | 95.3 | 84.6 | 84.7 |
| 4 | 63.6 | 97.0 | 113.3 | 71.7 | 137.2 | 118.4 | 106.9 | 119.4 | 103.4 | 118.3 | 131.2 | 112.6 | 118.3 | 106.2 | 92.0 | 95.5 |
| 5 | 101.4 | 127.7 | 107.6 | 121.6 | 140.8 | 114.1 | 130.3 | 117.3 | 140.4 | 141.8 | 150.2 | 121.0 | 136.7 | 148.9 | 100.9 | 100.7 |
| 6 | 127.7 | 203.9 | 117.7 | 134.6 | 141.0 | 120.8 | 106.6 | 145.5 | 154.8 | 142.6 | 150.2 | 154.7 | 161.3 | 151.7 | 116.8 | 86.5 |
| 7 | 81.0 | 225.2 | 199.6 | 129.9 | 160.2 | 157.2 | 237.9 | 204.5 | 198.5 | 110.9 | 156.6 | 151.0 | 201.8 | 171.5 | 109.3 | 83.4 |
| 8+ | 137.7 | 269.1 | 241.2 | 154.9 | 222.3 | 232.6 | 218.5 | 180.7 | 217.0 | 142.6 | 163.3 | 169.2 | 213.4 | 213.9 | 176.0 | 103.3 |
| Total | 46.1 | 30.9 | 32.2 | 40.7 | 39.5 | 41.9 | 24.5 | 47.4 | 49.0 | 33.6 | 29.7 | 35.8 | 39.5 | 37.3 | 28.7 | 34.3 |

[^1]Table 3.3.2 WESTERN BALTIC HERRING. Herring acoustic survey (HERAS) on the Spring Spawning Herring in the North Sea/Division IIIa in 1991-2007 (July).

| Year | 1991 | 1992* | 1993* | 1994* | 1995* | 1996* | 1997 | 1998 | 1999** | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Numbers in millions |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 |  | 3,853 | 372 | 964 |  |  |  |  |  |  |  |  |  |  |  |  |  | 112 |
| 1 |  | 277 | 103 | 5 | 2,199 | 1,091 | 128 | 138 | 1,367 | 1,509 | 66 | 3,346 | 1,833 | 1,669 | 2,687 | 2,081 | 3,918 | 5,852 |
| 2 | 1,864 | 2,092 | 2,768 | 413 | 1,887 | 1,005 | 715 | 1,682 | 1,143 | 1,891 | 641 | 1,577 | 1,110 | 930 | 1,342 | 2,217 | 3,621 | 1,160 |
| 3 | 1,927 | 1,799 | 1,274 | 935 | 1,022 | 247 | 787 | 901 | 523 | 674 | 452 | 1,393 | 395 | 726 | 464 | 1,780 | 933 | 843 |
| 4 | 866 | 1,593 | 598 | 501 | 1,270 | 141 | 166 | 282 | 135 | 364 | 153 | 524 | 323 | 307 | 201 | 490 | 499 | 333 |
| 5 | 350 | 556 | 434 | 239 | 255 | 119 | 67 | 111 | 28 | 186 | 96 | 88 | 103 | 184 | 103 | 180 | 154 | 274 |
| 6 | 88 | 197 | 154 | 186 | 174 | 37 | 69 | 51 | 3 | 56 | 38 | 40 | 25 | 72 | 84 | 27 | 34 | 176 |
| 7 | 72 | 122 | 63 | 62 | 39 | 20 | 80 | 31 | 2 | 7 | 23 | 18 | 12 | 22 | 37 | 10 | 26 | 45 |
| 8+ | 10 | 20 | 13 | 34 | 21 | 13 | 77 | 53 | 1 | 10 | 12 | 17 | 5 | 18 | 21 | 0.1 | 14 | 44 |
| Total | 5,177 | 10,509 | 5,779 | 3,339 | 6,867 | 2,673 | 2,088 | 3,248 | 3,201 | 4,696 | 1,481 | 7,002 | 3,807 | 3,926 | 4,939 | 6,786 | 9,199 | 8,839 |
| 3+ group | 5,177 | 4,287 | 2,536 | 1,957 | 2,781 | 577 | 1,245 | 1,428 | 691 | 1,295 | 774 | 2,079 | 864 | 1,328 | 910 | 2,487 | 1,660 | 1,715 |
| Biomass ('000 tonnnes) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 |  | 34.3 | 1 | 8.7 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 26.8 | 7 | 0.4 | 77.4 | 52.9 | 4.7 | 7.1 | 74.8 | 61.4 | 3.5 | 137.2 | 79.0 | 63.9 | 105.9 | 112.6 | 193.2 | 284.4 |
| 2 | 177.1 | 169.0 | 139 | 33.2 | 108.9 | 87.0 | 52.2 | 136.1 | 101.6 | 138.1 | 55.8 | 107.2 | 91.5 | 75.6 | 100.1 | 160.5 | 273.4 | 100.9 |
| 3 | 219.7 | 206.3 | 112 | 114.7 | 102.6 | 27.6 | 81.0 | 84.8 | 59.5 | 68.8 | 51.2 | 126.9 | 41.4 | 89.4 | 46.6 | 158.6 | 90.9 | 101.8 |
| 4 | 116.0 | 204.7 | 69 | 76.7 | 145.5 | 17.9 | 21.5 | 35.2 | 14.7 | 45.3 | 21.5 | 55.9 | 41.7 | 41.5 | 28.9 | 56.3 | 59.6 | 47.1 |
| 5 | 51.1 | 83.3 | 65 | 41.8 | 33.9 | 17.8 | 9.8 | 13.1 | 3.4 | 25.1 | 17.9 | 12.8 | 13.9 | 29.3 | 16.5 | 23.7 | 18.5 | 45.3 |
| 6 | 19.0 | 36.6 | 26 | 38.1 | 27.4 | 5.8 | 9.8 | 6.9 | 0.5 | 10.0 | 6.9 | 7.4 | 4.2 | 11.7 | 14.9 | 4.1 | 4.6 | 30.9 |
| 7 | 13.0 | 24.4 | 16 | 13.1 | 6.7 | 3.3 | 14.9 | 4.8 | 0.3 | 1.4 | 4.7 | 3.5 | 2.0 | 4.1 | 7.5 | 1.6 | 2.6 | 9.4 |
| 8+ | 2.0 | 5.0 | 2 | 7.8 | 3.8 | 2.7 | 13.6 | 9.0 | 0.1 | 1.3 | 2.7 | 3.1 | 0.9 | 3.2 | 4.9 | 0.02 | 1.94 | 8.65 |
| Total | 597.9 | 756.1 | 436.5 | 325.8 | 506.2 | 215.1 | 207.5 | 297.0 | 254.9 | 351.4 | 164.2 | 454.0 | 274.5 | 318.8 | 325.3 | 517.5 | 644.7 | 628.5 |
| $3+$ group | 420.9 | 560.3 | 291.0 | 292.3 | 319.9 | 75.2 | 150.6 | 153.7 | 78.5 | 151.9 | 104.9 | 209.6 | 104.0 | 179.3 | 119.3 | 244.4 | 178.2 | 243.2 |
| Mean weight (g) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| W-rings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 |  | 8.9 | 4.0 | 9.0 |  |  |  |  |  |  |  |  |  |  |  |  |  | 6.3 |
| 1 |  | 96.8 | 66.3 | 80.0 | 35.2 | 48.5 | 36.9 | 51.9 | 54.7 | 40.7 | 54.0 | 41.0 | 43.1 | 38.3 | 39.4 | 54.1 | 49.3 | 48.6 |
| 2 | 95.0 | 80.8 | 50.1 | 80.3 | 57.7 | 86.6 | 73.0 | 80.9 | 88.9 | 73.1 | 87.0 | 68.0 | 82.5 | 81.3 | 74.6 | 72.4 | 75.5 | 87.0 |
| 3 | 114.0 | 114.7 | 87.9 | 122.7 | 100.4 | 111.9 | 103.0 | 94.1 | 113.8 | 102.2 | 113.2 | 91.1 | 104.9 | 123.2 | 100.5 | 89.1 | 97.4 | 120.8 |
| 4 | 134.0 | 128.5 | 116.2 | 153.0 | 114.6 | 126.8 | 129.6 | 124.7 | 109.1 | 124.4 | 140.5 | 106.6 | 128.8 | 135.2 | 143.7 | 114.8 | 119.5 | 141.4 |
| 5 | 146.0 | 149.8 | 149.9 | 175.1 | 132.9 | 149.4 | 145.0 | 118.7 | 120.0 | 135.4 | 185.2 | 145.8 | 134.2 | 159.4 | 160.9 | 131.6 | 120.0 | 165.5 |
| 6 | 216.0 | 185.7 | 169.6 | 205.0 | 157.2 | 157.3 | 143.1 | 135.8 | 179.9 | 179.2 | 182.6 | 186.5 | 165.4 | 162.9 | 177.7 | 153.2 | 136.6 | 175.6 |
| 7 | 181.0 | 199.7 | 256.9 | 212.0 | 172.9 | 166.8 | 185.6 | 156.4 | 179.9 | 208.8 | 206.3 | 198.7 | 167.2 | 191.6 | 202.3 | 169.2 | 101.5 | 208.5 |
| 8+ | 200.0 | 252.0 | 164.2 | 230.3 | 183.1 | 212.9 | 178.0 | 168.0 | 181.7 | 135.2 | 226.9 | 183.4 | 170.3 | 178.0 | 229.2 | 178.0 | 138.3 | 196.7 |
| Total | 115.6 | 123.9 | 75.8 | 100.2 | 73.7 | 80.5 | 99.4 | 91.4 | 78.5 | 74.8 | 110.9 | 64.8 | 72.1 | 81.2 | 65.9 | 76.3 | 70.1 | 71.1 |

* revised in 1997
**the survey only covered the Skagerrak area by Norway. Additional estimates for the Kattegat area were added
(see ICES 2000/ACFM:10, Table 3.5.8)


## Table 3.3.3WESTERN BALTIC HERRING.

N20 Larval Abundance Index.
Estimation of 0-Group herring reaching 20 mm in length in Greifswalder Bodden and adjacent waters (March/April to June).

| Year | N20 <br> (millions) |
| :---: | :---: |
| $\mathbf{1 9 9 2}$ | 1,060 |
| 1993 | 3,044 |
| 1994 | 12,515 |
| 1995 | 7,930 |
| 1996 | 21,012 |
| 1997 | 4,872 |
| 1998 | 16,743 |
| 1999 | 20,364 |
| 2000 | 3,026 |
| 2001 | 4,845 |
| 2002 | 11,324 |
| 2003 | 5,507 |
| 2004 | 5,640 |
| 2005 | 3,887 |
| 2006 | 3,774 |
| 2007 | 1,900 |
| 2008 | 1,600 |

TABLE 3.6.1 WBSS HERRING. CATCH IN NUMBER

| Units $: ~ t h o u s a n d s ~$ |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 0 | 118958 | 145090 | 206102 | 263202 | 541302 | 171144 | 376795 | 549774 | 569599 | 152581 |
| 1 | 825969 | 456707 | 530707 | 249398 | 1660683 | 638877 | 668616 | 623072 | 616124 | 934545 |
| 2 | 541246 | 602624 | 495950 | 364980 | 438136 | 400585 | 289336 | 430903 | 334339 | 496396 |
| 3 | 564430 | 364864 | 415108 | 382650 | 226810 | 199681 | 276919 | 182860 | 246212 | 186615 |
| 4 | 279767 | 333993 | 260950 | 267033 | 194870 | 144155 | 75283 | 146685 | 90259 | 128625 |
| 5 | 177486 | 183200 | 210497 | 168142 | 84123 | 130086 | 43119 | 45322 | 55919 | 71727 |
| 6 | 46487 | 139835 | 102768 | 118416 | 60096 | 65274 | 39916 | 23759 | 15481 | 38262 |
| 7 | 13241 | 52660 | 63922 | 49504 | 32878 | 30705 | 21211 | 15400 | 9478 | 13777 |
| 8 | 4933 | 22574 | 24535 | 33088 | 20459 | 25111 | 24134 | 14112 | 6084 | 10689 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |  |  |
| 0 | 756285 | 150271 | 53489 | 243554 | 106906 | 7946 | 10721 | 9610 |  |  |
| 1 | 523163 | 659130 | 126876 | 457754 | 305171 | 148909 | 172044 | 150188 |  |  |
| 2 | 488816 | 281840 | 264855 | 197812 | 319225 | 187674 | 184735 | 138181 |  |  |
| 3 | 257837 | 321311 | 161251 | 164766 | 177833 | 233214 | 143904 | 135093 |  |  |
| 4 | 108097 | 172285 | 189432 | 93214 | 130394 | 150654 | 126861 | 88702 |  |  |
| 5 | 68376 | 57160 | 103648 | 91242 | 60639 | 98751 | 64996 | 85592 |  |  |
| 6 | 39092 | 38532 | 29117 | 48957 | 65695 | 42459 | 30199 | 44530 |  |  |
| 7 | 18307 | 13842 | 17452 | 14876 | 31231 | 32418 | 21256 | 17195 |  |  |
| 8 | 6687 | 8329 | 8819 | 11013 | 12620 | 17312 | 14759 | 18343 |  |  |

## TABLE 3.6.2 WBSS HERRING. WEIGHTS AT AGE IN THE CATCH

Units : kg
year
age $1991 \quad 1992 \quad 1993 \quad 1994 \quad 1995 \quad 1996 \quad 1997 \quad 1998 \quad 1999 \quad 2000$
00.02960 .01520 .01540 .01460 .01010 .01060 .02960 .01430 .01110 .0211
10.03480 .03450 .02540 .03700 .02090 .02460 .02750 .03330 .03430 .0255
$20.06690 .06730 .0680 \quad 0.08330 .06840 .08090 .06840 .06630 .06580 .0578$
30.09490 .09440 .1020 0.1032 0.0984 0.0970 0.1181 0.0942 0.0981 0.0950
40.12340 .11630 .11430 .12210 .12350 .11250 .13420 .11780 .11640 .1301
$\begin{array}{llllllllllllllll}5 & 0.1390 & 0.1417 & 0.1361 & 0.1411 & 0.1520 & 0.1328 & 0.1620 & 0.1367 & 0.1471 & 0.1428\end{array}$
60.15560 .16510 .16790 .15650 .17040 .13690 .18170 .16630 .15660 .1463
70.17090 .17580 .18230 .17050 .20630 .15420 .19670 .16520 .15380 .1583
80.18260 .19150 .19890 .18600 .21700 .19100 .20870 .18700 .15760 .1591 year
age $2001 \quad 2002 \quad 2003 \quad 2004 \quad 2005 \quad 2006 \quad 2007 \quad 2008$
00.01230 .01050 .01320 .006180 .01400 .01700 .01390 .0178
10.02430 .02130 .03150 .027540 .02720 .03600 .05060 .0644
20.05930 .07000 .06710 .064190 .07210 .07280 .07090 .0785
30.08620 .09680 .09070 .100170 .09380 .09820 .08540 .0957 40.10890 .11960 .10790 .105960 .11060 .11530 .11410 .1153 50.15670 .14000 .12230 .131390 .12280 .15350 .12880 .1399 60.15600 .18760 .13190 .152280 .14930 .15810 .15640 .1474 70.15560 .18140 .16030 .167680 .16190 .18650 .16730 .1680 80.17130 .17170 .16250 .152950 .17360 .18480 .19030 .1739

TABLE 3.6.3 WBSS HERRING. WEIGHTS AT AGE IN THE STOCK

```
Units : kg
    year
age 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000
    0 0.0001 0.0001 0.0001 0.0001 0.0001 0.0001 0.0001 0.0001 0.0001 0.0001
    1 0.0308 0.0203 0.0156 0.0186 0.0131 0.0181 0.0131 0.0221 0.0211 0.0140
    2 0.0528 0.0451 0.0402 0.0529 0.0459 0.0546 0.0515 0.0558 0.0567 0.0431
    3 0.0787 0.0818 0.0967 0.0836 0.0708 0.0905 0.1063 0.0829 0.0871 0.0837
    4 0.1041 0.1075 0.1079 0.1077 0.1327 0.1170 0.1333 0.1128 0.1081 0.1250
    5 0.1245 0.1313 0.1409 0.1392 0.1674 0.1197 0.1662 0.1338 0.1480 0.1436
    6 0.1449 0.1593 0.1671 0.1566 0.1892 0.1538 0.1943 0.1678 0.1601 0.1629
    7 0.1594 0.1710 0.1827 0.1768 0.2097 0.1467 0.2089 0.1683 0.1439 0.1650
    8 0.1640 0.1869 0.1891 0.2028 0.2338 0.1280 0.2263 0.1843 0.1504 0.1831
    year
age 2001 2002 2003 2004 2005 2006 2007 2008
    0 0.0001 0.0001 0.0001 0.0001 0.0001 0.0001 0.0001 0.0001
    1 0.0169 0.0164 0.0144 0.0131 0.0126 0.0185 0.0150 0.0290
    2 0.0509 0.0637 0.0445 0.0456 0.0514 0.0621 0.0550 0.0710
    3 0.0783 0.0905 0.0793 0.0811 0.0800 0.0953 0.0800 0.0910
    4 0.1159 0.1239 0.1051 0.1092 0.1066 0.1174 0.1140 0.1130
    5 0.1690 0.1736 0.1268 0.1440 0.1322 0.1659 0.1430 0.1480
    6 0.1763 0.1983 0.1506 0.1628 0.1573 0.1710 0.1710 0.1630
    7 0.1681 0.1980 0.1729 0.1932 0.1677 0.1858 0.1750 0.1850
    8 0.1805 0.2036 0.1847 0.2076 0.1820 0.1871 0.1880 0.1880
```

TABLE 3.6.4 WBSS HERRING. NATURAL MORTALITY

```
Units : NA
    year
age 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
    0}00.
    1 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5
```



```
    3 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 
    4 0.0.2
    5
    6
    7 0.2 0.2 0.2 0.2 0.2 0.2 0.2 
    8 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2
    year
age 2006 2007 2008
```



```
    1}0.5 0.5 0.
    0.2 0.2 0.2
    0.2 0.2 0.2
    4 0.2 0.2 0.2
    5 0.2 0.2 0.2
    0.2 0.2 0.2
    7 0.2 0.2 0.2
    0.2 0.2 0.2
```


## TABLE 3.6.5 WBSS HERRING. PROPORTION MATURE

```
Units : NA
    year
age 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
    0 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00
    1 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00
    2 0.20 0.20 0.20 0.20 0.20 0.20 0.20 0.20 0.20 0.20 0.20 0.20 0.20 0.20 0.20
    3 0.75 0.75 0.75 0.75 0.75 0.75 0.75 0.75 0.75 0.75 0.75 0.75 0.75 0.75 0.75
    4 0.90 0.90 0.90 0.90 0.90 0.90 0.90 0.90 0.90 0.90 0.90 0.90 0.90 0.90 0.90
    5 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    6 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    7 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    8 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    year
age 2006 2007 2008
    0 0.00 0.00 0.00
    10.00 0.00 0.00
    2 0.20 0.20 0.20
    3 0.75 0.75 0.75
    40.90 0.90 0.90
    5 1.00 1.00 1.00
    6 1.00 1.00 1.00
    71.00 1.00 1.00
    8 1.00 1.00 1.00
```

TABLE 3.6.6 WBSS HERRING. FRACTION OF HARVEST BEFORE SPAWNING

```
Units : NA
    year
age 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
    0 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 
    1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 
```



```
    3 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 
```




```
    6
```



```
    8 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 
    year
age 2006 2007 2008
    0}00.1 0.1 0.1
    1
    2 0.1 0.1 0.1
    3 0.1 0.1 0.1
```



```
    5 0.1 0.1 0.1
    6
```



```
    8 0.1 0.1 0.1
```

TABLE 3.6.7 WBSS HERRING. FRACTION OF NATURAL MORTALITY BEFORE SPAWNING

```
Units : NA
    year
age 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
    0 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25
    1 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25
    2 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25
    3 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25
    4 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25
    5 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25
    6 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25
    7 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25
    8 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0.25 0. 25 0.25
    year
age 2006 2007 2008
    0 0.25 0.25 0.25
    1 0.25 0.25 0.25
    2 0. 25 0.25 0.25
    30.25 0.25 0.25
    40.25 0.25 0.25
    5 0.25 0.25 0.25
    6 0.25 0.25 0.25
    70.25 0.25 0.25
    80.25 0.25 0.25
```


## TABLE 3.6.8 WBSS HERRING. SURVEY INDICES



## continued TABLE 3.6.8 WBSS HERRING. SURVEY INDICES

```
GerAS 1-3 wr - Configuration
\begin{tabular}{rrrrrr} 
min & max plusgroup & minyear & maxyear & startf & endf \\
1.00 & 3.00 & NA & 1994.00 & 2008.00 & 0.77
\end{tabular}
Index type : number
GerAS 1-3 wr - Index Values
Units : NA
    year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
    1 415730 1675340 1439460 1955400 801350 1338710 1429880 -1 837549 1238480
    2 883810
    3 559720 357960 434090 394530 394070 232510 328960 -1 575356 217270
        year
age 2004 2005 2006 2007 2008
    1 968860 752980 950450 560000 392780
    2 592360 640060 274460 278000 213500
    3446230 401070 376480 149000 209000
GerAS 1-3 wr - Index Variance (Inverse Weights)
Units : NA
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007 2008
\begin{tabular}{llllllllllllllll}
1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1
\end{tabular}
\begin{tabular}{llllllllllllllll}
2 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 \\
3 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1
\end{tabular}
N20 - Configuration
                min max plusgroup minyear maxyear startf endf
Index type : number
N20 - Index Values
Units : NA
        year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
    0 1060 3044 12515 7930 21012 4872 16743 20364 3026 4845 11324 5507 5640 3887
        year
age 2006 2007 2008
    0 3774 1900 1600
N20 - Index Variance (Inverse Weights)
Units : NA
    year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006
    year 1
age 2007 2008
```


## TABLE 3.6.9 WBSS HERRING. STOCK OBJECT CONFIGURATION

| min | max plusgroup | minyear | maxyear | minfbar | maxfbar |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 8 | 8 | 1991 | 2008 | 3 |

## TABLE 3.6.10 WBSS HERRING. FLICA CONFIGURATION SETTINGS

| sep. 2 | NA |
| :---: | :---: |
| sep.gradual | TRUE |
| sr | FALSE |
| sr.age | 0 |
| lambda.age | 0.1111111110 |
| lambda.yr | 11111 |
| lambda.sr | 0 |
| index.model | linear linear linear |
| index.cor | 111 |
| sep.nyr | 5 |
| sep.age | 4 |
| sep.sel | 1 |

## TABLE 3.6.11 WBSS HERRING. FLR, R SOFTWARE VERSIONS

R version 2.8.1 (2008-12-22)

```
Package : FLICA
Version : 1.4-10
Packaged : Sat Mar 21 18:30:56 2009; mpa
Built : R 2.8.0; ; 2009-03-21 18:30:58; windows
Package : FLAssess
Version : 1.99-102
Packaged : Mon Mar 23 08:18:19 2009; mpa
Built : R 2.8.0; i386-pc-mingw32; 2009-03-23 08:18:21; windows
Package : FLCore
Version : 3.0
Packaged : Fri Apr 3 09:33:49 2009; mpa
Built : R 2.8.0; i386-pc-mingw32; 2009-04-03 09:33:52; windows
```

TABLE 3.6.12 WBSS HERRING. STOCK SUMMARY

|  | Recr | TSB | SSB | Fbar | Landings | Landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 0 |  |  | (Ages 3-6) |  | SOP |
|  |  |  |  | $f$ | tonnes |  |
| 1991 | 5003979 | 616744 | 310543 | 0.358 | 191573 | 1.000 |
| 1992 | 3652584 | 541902 | 322123 | 0.476 | 194411 | 1.000 |
| 1993 | 3109372 | 464010 | 295303 | 0.544 | 185010 | 1.000 |
| 1994 | 6182874 | 377101 | 231273 | 0.689 | 172438 | 1.000 |
| 1995 | 4050991 | 318714 | 183178 | 0.512 | 150831 | 1.000 |
| 1996 | 4472243 | 273456 | 134412 | 0.701 | 121266 | 1.000 |
| 1997 | 3991042 | 275540 | 150554 | 0.510 | 115588 | 1.000 |
| 1998 | 5585578 | 272681 | 121510 | 0.492 | 107032 | 1.000 |
| 1999 | 6439794 | 289625 | 128418 | 0.373 | 97240 | 1.000 |
| 2000 | 3444329 | 294912 | 141946 | 0.466 | 109914 | 1.000 |
| 2001 | 4448602 | 319603 | 162715 | 0.454 | 105803 | 1.000 |
| 2002 | 3057143 | 351641 | 201961 | 0.410 | 106191 | 1.000 |
| 2003 | 4039490 | 266620 | 162065 | 0.394 | 78309 | 1.000 |
| 2004 | 2655128 | 281995 | 167171 | 0.358 | 76815 | 1.000 |
| 2005 | 2226088 | 286039 | 165281 | 0.407 | 88406 | 1.000 |
| 2006 | 1813432 | 314088 | 192109 | 0.396 | 90549 | 1.000 |
| 2007 | 1259682 | 248883 | 161537 | 0.358 | 68997 | 0.988 |
| 2008 | 894443 | 253761 | 159406 | 0.367 | 68484 | 1.000 |

TABLE 3.6.13 WBSS HERRING. ESTIMATED FISHING MORTALITY

```
Units : f
    year
age 1991 1992 1993 1994 1995 1996 1997 1998
    0 0.0279 0.047 0.0797 0.0505 0.167 0.0453 0.115 0.121 0.108 0.0526 0.218
    1 0.2591 0.174 0.2984 0.1602 0.639 0.3781 0.308 0.351 0.237 0.3197 0.316
    2 0.3195 0.372 0.3512 0.4219 0.572 0.3783 0.358 0.407 0.394 0.3714 0. 0.334
    3 0.4206 0.371 0.4747 0.5033 0.507 0.5617 0.490 0.403 0.432 0.3998 0.336
    4 0.3984 0.474 0.4961 0.6462 0.522 0.7156 0.427 0.526 0.356 0.4225 0.427
    5 0.3714 0.496 0.6269 0.7007 0.432 0.8130 0.483 0.497 0.390 0.5344 0.418
    6 0.2399 0.565 0.5780 0.9072 0.587 0.7125 0.638 0.541 0.314 0.5076 0.634
    7 0.4106 0.468 0.5511 0.6163 0.699 0.6883 0.534 0.547 0.432 0.5103 0.489
    8 0.4106 0.468 0.5511 0.6163 0.699 0.6883 0.534 0.547 0.432 0.5103 0.489
        year
age 2002 2003 2004 2005 2006 2007 2008
    0 0.0585 0.0154 0.0199 0.0227 0.0220 0.0199 0.0204
    1 0.3728 0.0782 0.1833 0.2087 0.2029 0.1833 0.1881
    2 0.3409 0.3046 0.2424 0.2760 0.2683 0.2423 0.2487
    3 0.3834 0.3339 0.3014 0.3432 0.3336 0.3013 0.3092
    4 0.3943 0.4100 0.3608 0.4108 0.3993 0.3607 0.3701
    5 0.4218 0.4384 0.3761 0.4282 0.4162 0.3759 0.3858
    6 0.4411 0.3954 0.3926 0.4470 0.4345 0.3925 0.4028
    7 0.4840 0.3668 0.3608 0.4108 0.3993 0.3607 0.3701
    8 0.4840 0.3668 0.3608 0.4108 0.3993 0.3607 0.3701
```

TABLE 3.6.14 WBSS HERRING. ESTIMATED POPULATION ABUNDANCE

| Units <br> year <br> age <br> 0 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 5003979 | 3652584 | 3109372 | 6182874 | 4050991 | 4472243 | 3991042 | 5585578 | 6439794 |
| 2 | 2171998 | 3605104 | 2581634 | 2127100 | 4354977 | 2538818 | 3166531 | 2634467 | 3667888 |
| 3 | 1801225 | 1291944 | 1837128 | 1161808 | 1099199 | 1394465 | 1055070 | 1411769 | 1124560 |
| 4 | 933200 | 968392 | 730181 | 1058709 | 623807 | 507889 | 782094 | 603999 | 769229 |
| 5 | 627394 | 512981 | 493505 | 364024 | 524004 | 307557 | 237112 | 392196 | 330430 |
| 6 | 239518 | 354323 | 255855 | 215856 | 14789 | 254500 | 123111 | 126606 | 189745 |
| 7 | 43085 | 154275 | 164950 | 117524 | 71335 | 67321 | 92414 | 62155 | 63051 |
| 8 | 16052 | 66134 | 63312 | 78552 | 44389 | 55056 | 63818 | 39982 | 29617 |
| year |  |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| 0 | 3444329 | 4448602 | 3057143 | 4039490 | 2655128 | 2226088 | 1813432 | 1259682 | 894443 |
| 1 | 4283602 | 2420957 | 2650607 | 2136119 | 2946679 | 1928203 | 1612172 | 1314150 | 914810 |
| 2 | 1754449 | 1887276 | 1070384 | 1107396 | 1198149 | 1487878 | 949184 | 798253 | 663594 |
| 3 | 620661 | 990765 | 1106052 | 623198 | 668615 | 769803 | 924364 | 594256 | 512908 |
| 4 | 408966 | 340693 | 579554 | 617148 | 365373 | 404973 | 447183 | 542139 | 359966 |
| 5 | 189480 | 219453 | 181977 | 319882 | 335317 | 208547 | 219874 | 245593 | 309473 |
| 6 | 105163 | 90909 | 118333 | 97715 | 168947 | 188486 | 111272 | 118727 | 138065 |
| 7 | 37710 | 51826 | 39485 | 62329 | 53872 | 93410 | 98693 | 58995 | 65651 |
| 8 | 29257 | 18930 | 23759 | 31497 | 39883 | 41055 | 57638 | 53461 | 65021 |

TABLE 3.6.15 WBSS HERRING. SURVIVORS AFTER TERMINAL YEAR

| Units |  |
| :--- | ---: |
| year |  |
| age | 2009 |
| 0 | NA |
| 1 | 649225 |
| 2 | 459723 |
| 3 | 423677 |
| 4 | 308239 |
| 5 | 203543 |
| 6 | 172266 |
| 7 | 75561 |
| 8 | 73889 |

TABLE 3.6.16 WBSS HERRING. FITTED SELECTION PATTERN

```
Units : NA
    year
age 2004 2005 2006 2007 2008
    0 0.0552 0.0552 0.0552 0.0552 0.0552
    1 0.5082 0.5082 0.5082 0.5082 0.5082
    2 0.6719 0.6719 0.6719 0.6719 0.6719
    3 0.8354 0.8354 0.8354 0.8354 0.8354
    4 1.0000 1.0000 1.0000 1.0000 1.0000
    5 1.0424 1.0424 1.0424 1.0424 1.0424
    6 1.0882 1.0882 1.0882 1.0882 1.0882
    7 1.0000 1.0000 1.0000 1.0000 1.0000
    8 1.0000 1.0000 1.0000 1.0000 1.0000
```


## TABLE 3.6.17 WBSS HERRING. PREDICTED CATCH IN NUMBERS

| Uni | $\begin{aligned} & \text { ts : } \\ & \text { year } \end{aligned}$ | A |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| 0 | 118958 | 145090 | 206102 | 263202 | 541302 | 171144 | 376795 | 549774 | 569599 | 152581 |
| 1 | 825969 | 456707 | 530707 | 249398 | 1660683 | 638877 | 668616 | 623072 | 616124 | 934545 |
| 2 | 541246 | 602624 | 495950 | 364980 | 438136 | 400585 | 289336 | 430903 | 334339 | 496396 |
| 3 | 564430 | 364864 | 415108 | 382650 | 226810 | 199681 | 276919 | 182860 | 246212 | 186615 |
| 4 | 279767 | 333993 | 260950 | 267033 | 194870 | 144155 | 75283 | 146685 | 90259 | 128625 |
| 5 | 177486 | 183200 | 210497 | 168142 | 84123 | 130086 | 43119 | 45322 | 55919 | 71727 |
| 6 | 46487 | 139835 | 102768 | 118416 | 60096 | 65274 | 39916 | 23759 | 15481 | 38262 |
| 7 | 13241 | 52660 | 63922 | 49504 | 32878 | 30705 | 21211 | 15400 | 9478 | 13777 |
| 8 | 4933 | 22574 | 24535 | 33088 | 20459 | 25111 | 24134 | 14112 | 6084 | 10689 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |  |  |
| 0 | 756285 | 150271 | 53489 | 45230 | 43121 | 34157 | 21452 | 15629 |  |  |
| 1 | 523163 | 659130 | 126876 | 391377 | 288344 | 234957 | 174499 | 124399 |  |  |
| 2 | 488816 | 281840 | 264855 | 234699 | 326743 | 203345 | 156326 | 132978 |  |  |
| 3 | 257837 | 321311 | 161251 | 158477 | 203827 | 238958 | 140817 | 124284 |  |  |
| 4 | 108097 | 172285 | 189432 | 100892 | 124486 | 134315 | 149668 | 101550 |  |  |
| 5 | 68376 | 57160 | 103648 | 95853 | 66304 | 68319 | 70188 | 90365 |  |  |
| 6 | 39092 | 38532 | 29117 | 50043 | 62036 | 35800 | 35159 | 41766 |  |  |
| 7 | 18307 | 13842 | 17452 | 14876 | 28714 | 29643 | 16287 | 18521 |  |  |
| 8 | 6687 | 8329 | 8819 | 11013 | 12620 | 17312 | 14759 | 18343 |  |  |

TABLE 3.6.18 WBSS HERRING. CATCH RESIDUALS

| Units <br> year |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
| age thousands NA |  |  |  |  |  |
| ar | 2004 | 2005 | 2006 | 2007 | 2008 |
| 0 | 1.6836 | 0.9079 | -1.4583 | -0.6936 | -0.4863 |
| 1 | 0.1567 | 0.0567 | -0.4561 | -0.0142 | 0.1884 |
| 2 | -0.1710 | -0.0233 | -0.0802 | 0.1670 | 0.0384 |
| 3 | 0.0389 | -0.1364 | -0.0243 | 0.0217 | 0.0834 |
| 4 | -0.0792 | 0.0464 | 0.1148 | -0.1653 | -0.1353 |
| 5 | -0.0493 | -0.0893 | 0.3684 | -0.0769 | -0.0543 |
| 6 | -0.0219 | 0.0573 | 0.1706 | -0.1521 | 0.0641 |
| 7 | 0.0000 | 0.0840 | 0.0895 | 0.2663 | -0.0743 |
| 8 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |

TABLE 3.6.19 WBSS HERRING. PREDICTED INDEX VALUES

```
HERAS 3-6 wr
Units : NA NA
        year
age 1993 1994 1995 1906 1906 1997 1998 1999
        3 1174825597 1016025257 597208957 469935736 756729003 617080572 NA
        4 561793112 429007990 396633155 206296485 190425267 296144842 NA
        5 286302391 201663969 172205693 131433623 78121659 79655213 NA
        6 124664870 85613207 71654004 62574134 43374240
        year 2000 2001 2002 2002 2003 2004
        3 635442343 1055484031 1144031766 664850931 727965953 816534810 986368646
        4 329461198 273671585 475184947 501059799 305923104 328645320 365510811
        5 116467534 145104712 120011069 208786922 227553562 136987833 145511751
        6 53543422 42772298 62807435 53364867 92431100 99672601 59302405
        year
    age 2007 2008
        3647041263555710453
        4453956297299634253
        5 166675711 208736005
        6 64960551 75055786
GerAS 1-3 wr
Units : NA NA
        year
        age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
        1 925227 1291674 927643 1223932 983454 1499773 1640061 NA 972616 992118
        2 411636 345363 511605 393488 506108 407311 647219 NA 404627 430970
        3460871 270712 210995 344041 284857 354582 293506 NA 529935 310650
        year
    age 2004 2005 2006 2007 2008
        1 1258201 806753 677679 561152 389128
    2 490063 592426 380272 326517 270057
    3 342083 380908 460907 304060 260780
```


## continued TABLE 3.6.19 WBSS HERRING. PREDICTED INDEX VALUES

```
N20
Units : NA NA
    year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
    0 6158 5174 10409 6509 7545 6547 9144 10596 5794 7005 5130 6897 4525 3790
    year
age 2006 2007 2008
    0 3088 2147 1524
```

TABLE 3.6.20 WBSS HERRING. INDEX RESIDUALS

```
HERAS 3-6 wr
Units : NA
    year
age 1993 1904 1995 1996 1997 1998 1999 2000 2001 2002
    3 0.0810 -0.0831 0.537 -0.6432 0.0392 0.3785 NA 0.0583 -0.847 0.1968
    4 0.0625 0.1551 1.164 -0.3806 -0.1373-0.0489 NA 0.0994 -0.581 0.0984
    5 0.4160 0.1699 0.393-0.0994 -0.1536 0.3318 NA 0.4665 -0.409 -0.3159
    6 0.2113 0.7759 0.887 -0.5254 0.4642 0.4983 NA 0.0377 -0.129 -0.4638
        year
age 2003 2004 2005 2006 2007 2008
    3-0.522 -0.00270 -0.566 0.591 0.3660 0.417
    4-0.438 0.00319-0.490 0.293 0.0946 0.106
    5 -0.703 -0.21408-0.290 0.215 -0.0791 0.272
    6-0.750 -0.24841-0.176 -0.787 -0.6474 0.852
GerAS 1-3 wr
Units : NA
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004
    1 -0.800 0.2601 0.439 0.469 -0.205 -0.114 -0.137 NA -0.1495 0.222 -0.2613
    2 0.764 -0.0497 0.143 0.629 0.293 -0.349 -0.355 NA 0.0406 -0.661 0.1896
    3 0.194 0.2794 0.721 0.137 0.325 -0.422 0.114 NA 0.0822 -0.358 0.0120
        year
age 2005 2006 2007 2008
    1 -0.0690 0.338-0.00206 0.00934
    2 0.0773 -0.326 -0.16086 -0.23499
    3 0.0516 -0.202 -0.71328 -0.22134
N20
Units : NA
        year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 
    0 -1.76 -0.53 0.184 0.198 1.02 -0.295 0.605 0.653-0.65 -0.369 0.792 -0. 225
    year
age 2004 2005 2006 2007 2008
    0 0.220 0.0254 0.201 -0.122 0.0486
```


## TABLE 3.6.21 WBSS HERRING. FIT PARAMETERS

|  | Value | Std.dev | Lower.95.pct.CL | Upper.95.pct.CL |
| :---: | :---: | :---: | :---: | :---: |
| F, 2004 | 0.36075 | 0.1975 | 0.24498 | 0.53124 |
| F, 2005 | 0.41076 | 0.1949 | 0.28036 | 0.60181 |
| F, 2006 | 0.39929 | 0.1996 | 0.27000 | 0.59049 |
| F, 2007 | 0.36065 | 0.2114 | 0.23832 | 0.54577 |
| F, 2008 | 0.37013 | 0.2293 | 0.23613 | 0.58017 |
| Selectivity at age 0 | 0.05517 | 0.4815 | 0.02147 | 0.14178 |
| Selectivity at age 1 | 0.50818 | 0.2184 | 0.33118 | 0.77976 |
| Selectivity at age 2 | 0.67192 | 0.2107 | 0.44456 | 1.01556 |
| Selectivity at age 3 | 0.83543 | 0.2035 | 0.56061 | 1.24498 |
| Selectivity at age 5 | 1.04241 | 0.1807 | 0.73147 | 1.48553 |
| Selectivity at age 6 | 1.08823 | 0.1740 | 0.77374 | 1.53055 |
| Terminal year pop, age 0 | 894441.84149 | 0.3146 | 482824.28584 | 1656971.76234 |
| Terminal year pop, age 1 | 914808.50544 | 0.2301 | 582750.44664 | 1436077.15180 |
| Terminal year pop, age 2 | 663592.81555 | 0.1994 | 448941.82846 | 980874.12877 |
| Terminal year pop, age 3 | 512906.88600 | 0.1773 | 362347.58894 | 726025.18061 |
| Terminal year pop, age 4 | 359964.57231 | 0.1738 | 256027. 26947 | 506096.45443 |
| Terminal year pop, age 5 | 309471.84907 | 0.1873 | 214382.38181 | 446738. 32130 |
| Terminal year pop, age 6 | 138064.40500 | 0.2085 | 91741.74668 | 207776.50979 |
| Terminal year pop, age 7 | 65650.18409 | 0.2455 | 40576.13612 | 106218.75524 |
| Last true age pop, 2004 | 53871.15392 | 0.3567 | 26775.01885 | 108388.39149 |
| Last true age pop, 2005 | 93408.84018 | 0.2694 | 55090.80818 | 158378.71529 |
| Last true age pop, 2006 | 98692.24358 | 0.2446 | 61100.16492 | 159412.97305 |
| Last true age pop, 2007 | 58994. 19500 | 0.2490 | 36210.72573 | 96112.82220 |
| Index 1, age 3 numbers, Q | 1489.46015 | 0.1665 | 1074.81391 | 2064.07038 |
| Index 1, age 4 numbers, Q | 1188.73064 | 0.1674 | 856.22207 | 1650.36688 |
| Index 1, age 5 numbers, Q | 972.72139 | 0.1695 | 697.80467 | 1355.94810 |
| Index 1, age 6 numbers, Q | 792.34929 | 0.1735 | 563.90278 | 1113.34332 |
| Index 2, age 1 numbers, Q | 0.73761 | 0.1511 | 0.54857 | 0.99181 |
| Index 2, age 2 numbers, Q | 0.58270 | 0.1508 | 0.43362 | 0.78304 |
| Index 2, age 3 numbers, Q | 0.76411 | 0.1509 | 0.56850 | 1.02703 |
| Index 3, age 0 numbers, Q | 0.00194 | 0.0833 | 0.00165 | 0.00228 |

Table 3.7.1 WESTERN BALTIC HERRING. Parameters used for short term prediction and single option tables.

2009 (Intermediate Year)

| Age | $\mathbf{N}$ | $\mathbf{M}$ | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 2225068 | 0.30 | 0.00 | 0.10 | 0.25 | 0.000 | 0.021 | 0.016 |
| 1 | 649225 | 0.50 | 0.00 | 0.10 | 0.25 | 0.021 | 0.191 | 0.050 |
| 2 | 459723 | 0.20 | 0.20 | 0.10 | 0.25 | 0.063 | 0.253 | 0.074 |
| 3 | 423677 | 0.20 | 0.75 | 0.10 | 0.25 | 0.089 | 0.315 | 0.093 |
| 4 | 308239 | 0.20 | 0.90 | 0.10 | 0.25 | 0.115 | 0.377 | 0.115 |
| 5 | 203543 | 0.20 | 1.00 | 0.10 | 0.25 | 0.152 | 0.393 | 0.141 |
| 6 | 172266 | 0.20 | 1.00 | 0.10 | 0.25 | 0.168 | 0.410 | 0.154 |
| 7 | 75561 | 0.20 | 1.00 | 0.10 | 0.25 | 0.182 | 0.377 | 0.174 |
| 8 | 73889 | 0.20 | 1.00 | 0.10 | 0.25 | 0.188 | 0.377 | 0.183 |

2010 (Advice Year)

| Age | $\mathbf{N}$ | $\mathbf{M}$ | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 2225068 | 0.30 | 0.00 | 0.10 | 0.25 | 0.000 | 0.021 | 0.016 |
| 1 | - | 0.50 | 0.00 | 0.10 | 0.25 | 0.021 | 0.191 | 0.050 |
| 2 | - | 0.20 | 0.20 | 0.10 | 0.25 | 0.063 | 0.253 | 0.074 |
| 3 | - | 0.20 | 0.75 | 0.10 | 0.25 | 0.089 | 0.315 | 0.093 |
| 4 | - | 0.20 | 0.90 | 0.10 | 0.25 | 0.115 | 0.377 | 0.115 |
| 5 | - | 0.20 | 1.00 | 0.10 | 0.25 | 0.152 | 0.393 | 0.141 |
| 6 | - | 0.20 | 1.00 | 0.10 | 0.25 | 0.168 | 0.410 | 0.154 |
| 7 | - | 0.20 | 1.00 | 0.10 | 0.25 | 0.182 | 0.377 | 0.174 |
| 8 | - | 0.20 | 1.00 | 0.10 | 0.25 | 0.188 | 0.377 | 0.183 |

2011 (Continuation Year)

| Age | $\mathbf{N}$ | $\mathbf{M}$ | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 2225068 | 0.30 | 0.00 | 0.10 | 0.25 | 0.000 | 0.021 | 0.016 |
| 1 | - | 0.50 | 0.00 | 0.10 | 0.25 | 0.021 | 0.191 | 0.050 |
| 2 | - | 0.20 | 0.20 | 0.10 | 0.25 | 0.063 | 0.253 | 0.074 |
| 3 | - | 0.20 | 0.75 | 0.10 | 0.25 | 0.089 | 0.315 | 0.093 |
| 4 | - | 0.20 | 0.90 | 0.10 | 0.25 | 0.115 | 0.377 | 0.115 |
| 5 | - | 0.20 | 1.00 | 0.10 | 0.25 | 0.152 | 0.393 | 0.141 |
| 6 | - | 0.20 | 1.00 | 0.10 | 0.25 | 0.168 | 0.410 | 0.154 |
| 7 | - | 0.20 | 1.00 | 0.10 | 0.25 | 0.182 | 0.377 | 0.174 |
| 8 | - | 0.20 | 1.00 | 0.10 | 0.25 | 0.188 | 0.377 | 0.183 |


| Age | Age (in winter rings) | SWt | Weight in the stock (kg) |
| :--- | :--- | :--- | :--- |
| N | Stock numbers (thousands) | Sel | Exploitation pattern |
| M | Natural mortality $\left(. \mathrm{yr}^{-1}\right)$ | CWt | Weight in the catch (kg) |
| Mat | Maturity ogive |  |  |
| PF | Proportion of fishing mortality (F) before spawning |  |  |
| PM | Proportion of natural mortality (M) before spawning |  |  |

$\mathrm{N}_{2009,2010,} 2011$ age 0
$\mathrm{N}_{2009}$ age 1-8+
Natural Mortality (M)
Weight in the Catch/Stock (CWt/SWt)
Selection pattern (Sel)

Geometric mean recruitment (Table 3.6.14) from 2003-2007
Output from assessment (Table 3.6.15)
Arithmetic mean from 2006-2008
Arithmetic mean from 2006-2008
Arithmetic mean from 2006-2008

Table 3.7.2 WESTERN BALTIC HERRING. Short term prediction single option tables. Left hand tables) stock numbers ( N ) and fishing mortality ( F ) in each year. Right hand tables) Spawning stock biomass (SSB), mean fishing mortality (F.bar) and total catch of WBSS herring (Yield).

| a). Catch(2010) $=$ Zero |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | $\mathbf{N ( 2 0 0 9 )}$ | $\mathbf{N ( 2 0 1 0 )}$ | $\mathbf{N ( 2 0 1 1 )}$ | $\mathbf{F}(\mathbf{2 0 0 9 )}$ | $\mathbf{F ( 2 0 1 0 )}$ | $\mathbf{F ( 2 0 1 1 )}$ |
| 0 | 2225068 | 2225068 | 2225068 | 0.0161 | 0.0000 | 0.0000 |
| 1 | 649225 | 1622036 | 1648371 | 0.1483 | 0.0000 | 0.0000 |
| 2 | 459723 | 339491 | 983815 | 0.1961 | 0.0000 | 0.0000 |
| 3 | 423677 | 309357 | 277951 | 0.2439 | 0.0000 | 0.0000 |
| 4 | 308239 | 271813 | 253280 | 0.2919 | 0.0000 | 0.0000 |
| 5 | 203543 | 188478 | 222542 | 0.3043 | 0.0000 | 0.0000 |
| 6 | 172266 | 122929 | 154313 | 0.3176 | 0.0000 | 0.0000 |
| 7 | 75561 | 102657 | 100646 | 0.2919 | 0.0000 | 0.0000 |
| 8 | 73889 | 91384 | 158867 | 0.2919 | 0.0000 | 0.0000 |


| Age | N(2009) | N(2010) | N(2011) | F(2009) | F(2010) | F(2011) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2225068 | 2225068 | 2225068 | 0.0161 | 0.0172 | 0.0172 |
| 1 | 649225 | 1622036 | 1620224 | 0.1483 | 0.1586 | 0.1586 |
| 2 | 459723 | 339491 | 839498 | 0.1961 | 0.2097 | 0.2097 |
| 3 | 423677 | 309357 | 225360 | 0.2439 | 0.2608 | 0.2608 |
| 4 | 308239 | 271813 | 195138 | 0.2919 | 0.3122 | 0.3122 |
| 5 | 203543 | 188478 | 162870 | 0.3043 | 0.3254 | 0.3254 |
| 6 | 172266 | 122929 | 111451 | 0.3176 | 0.3397 | 0.3397 |
| 7 | 75561 | 102657 | 71658 | 0.2919 | 0.3122 | 0.3122 |
| 8 | 73889 | 91384 | 116269 | 0.2919 | 0.3122 | 0.3122 |

c). $\operatorname{Catch}(2010)=2009$ TACs (56 627 t )

| Age | $\mathbf{N ( 2 0 0 9 )}$ | $\mathbf{N ( 2 0 1 0 )}$ | $\mathbf{N ( 2 0 1 1 )}$ | $\mathbf{F}(\mathbf{2 0 0 9 )}$ | $\mathbf{F ( 2 0 1 0 )}$ | $\mathbf{F ( 2 0 1 1 )}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 2225068 | 2225068 | 2225068 | 0.0161 | 0.0208 | 0.0208 |
| 1 | 649225 | 1622036 | 1614476 | 0.1483 | 0.1914 | 0.1914 |
| 2 | 459723 | 339491 | 812465 | 0.1961 | 0.2530 | 0.2530 |
| 3 | 423677 | 309357 | 215815 | 0.2439 | 0.3146 | 0.3146 |
| 4 | 308239 | 271813 | 184915 | 0.2919 | 0.3766 | 0.3766 |
| 5 | 203543 | 188478 | 152710 | 0.3043 | 0.3925 | 0.3925 |
| 6 | 172266 | 122929 | 104214 | 0.3176 | 0.4098 | 0.4098 |
| 7 | 75561 | 102657 | 66807 | 0.2919 | 0.3766 | 0.3766 |
| 8 | 73889 | 91384 | 109016 | 0.2919 | 0.3766 | 0.3766 |

d). $\operatorname{Catch}(2010)=2009$ TACs $+15 \%$ ( 65121 t)

| Age | $\mathbf{N ( 2 0 0 9 )}$ | $\mathbf{N}(\mathbf{2 0 1 0})$ | $\mathbf{N ( 2 0 1 1 )}$ | $\mathrm{F}(\mathbf{2 0 0 9 )}$ | $\mathrm{F}(\mathbf{2 0 1 0 )}$ | $\mathrm{F}(\mathbf{2 0 1 1 )}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 2225068 | 2225068 | 2225068 | 0.0161 | 0.0245 | 0.0245 |
| 1 | 649225 | 1622036 | 1608435 | 0.1483 | 0.2259 | 0.2259 |
| 2 | 459723 | 339491 | 784892 | 0.1961 | 0.2987 | 0.2987 |
| 3 | 423677 | 309357 | 206184 | 0.2439 | 0.3714 | 0.3714 |
| 4 | 308239 | 271813 | 174711 | 0.2919 | 0.4445 | 0.4445 |
| 5 | 203543 | 188478 | 142680 | 0.3043 | 0.4634 | 0.4634 |
| 6 | 172266 | 122929 | 97088 | 0.3176 | 0.4837 | 0.4837 |
| 7 | 75561 | 102657 | 62046 | 0.2919 | 0.4445 | 0.4445 |
| 8 | 73889 | 91384 | 101856 | 0.2919 | 0.4445 | 0.4445 |

e). $\operatorname{Catch}(2010)=2009 \operatorname{Catch}(45087 \mathrm{t})$

| Age | $\mathbf{N}(\mathbf{2 0 0 9})$ | $\mathbf{N ( 2 0 1 0 )}$ | $\mathbf{N}(\mathbf{2 0 1 1})$ | $\mathbf{F}(\mathbf{2 0 0 9 )}$ | $\mathbf{F}(\mathbf{2 0 1 0})$ | $\mathbf{F ( 2 0 1 1 )}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 2225068 | 2225068 | 2225068 | 0.0161 | 0.0160 | 0.0154 |
| 1 | 649225 | 1622036 | 1622218 | 0.1483 | 0.1473 | 0.1416 |
| 2 | 459723 | 339491 | 849067 | 0.1961 | 0.1948 | 0.1873 |
| 3 | 423677 | 309357 | 228763 | 0.2439 | 0.2422 | 0.2329 |
| 4 | 308239 | 271813 | 198808 | 0.2919 | 0.2899 | 0.2787 |
| 5 | 203543 | 188478 | 166543 | 0.3043 | 0.3022 | 0.2905 |
| 6 | 172266 | 122929 | 114072 | 0.3176 | 0.3154 | 0.3033 |
| 7 | 75561 | 102657 | 73418 | 0.2919 | 0.2899 | 0.2787 |
| 8 | 73889 | 91384 | 118891 | 0.2919 | 0.2899 | 0.2787 |

f). $\operatorname{Fbar}(2010)=0.25$

| Age | $\mathbf{N}(\mathbf{2 0 0 9 )}$ | $\mathbf{N ( 2 0 1 0 )}$ | $\mathbf{N ( 2 0 1 1 )}$ | $\mathbf{F ( 2 0 0 9 )}$ | $\mathbf{F ( 2 0 1 0 )}$ | $\mathbf{F ( 2 0 1 1 )}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 2225068 | 2225068 | 2225068 | 0.0161 | 0.0139 | 0.0139 |
| 1 | 649225 | 1622036 | 1625598 | 0.1483 | 0.1281 | 0.1281 |
| 2 | 459723 | 339491 | 865499 | 0.1961 | 0.1694 | 0.1694 |
| 3 | 423677 | 309357 | 234635 | 0.2439 | 0.2106 | 0.2106 |
| 4 | 308239 | 271813 | 205172 | 0.2919 | 0.2521 | 0.2521 |
| 5 | 203543 | 188478 | 172946 | 0.3043 | 0.2628 | 0.2628 |
| 6 | 172266 | 122929 | 118647 | 0.3176 | 0.2744 | 0.2744 |
| 7 | 75561 | 102657 | 76495 | 0.2919 | 0.2521 | 0.2521 |
| 8 | 73889 | 91384 | 123462 | 0.2919 | 0.2521 | 0.2521 |


| Year | SSB | F.bar | Yield |
| ---: | ---: | ---: | ---: |
| 2009 | 141824 | 0.2894 | 45087 |
| 2010 | 128196 | 0.2500 | 39808 |
| 2011 | 121656 | 0.2500 | 42254 |

Table 3.7.3 WESTERN BALTIC HERRING. Short-term prediction multiple option table, based on a catch constraint in the intermediate year of 45087 t .

| 2009 <br> Fmult | Fbar | Landings | SSB | $\mathbf{2 0 1 0}$ <br> Fmult | Fbar | Landings | SSB | $\mathbf{2 0 1 1}$ <br> SSB |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0.789 | 0.289 | 45087 | 141824 | 0.000 | 0.000 | 0 | 131427 | 156964 |
|  | - | - | - | 0.100 | 0.029 | 5044 | 131049 | 152385 |
|  | - | - | - | 0.200 | 0.058 | 9967 | 130672 | 147943 |
|  | - | - | - | 0.300 | 0.087 | 14772 | 130295 | 143635 |
|  | - | - | - | 0.400 | 0.116 | 19463 | 129921 | 139457 |
|  | - | - | - | 0.500 | 0.145 | 24043 | 129547 | 135403 |
|  | - | - | - | 0.600 | 0.174 | 28514 | 129174 | 131472 |
|  | - | - | - | 0.700 | 0.203 | 32879 | 128802 | 127658 |
|  | - | - | - | 0.800 | 0.232 | 37141 | 128432 | 123958 |
|  | - | - | - | 0.900 | 0.260 | 41303 | 128062 | 120369 |
|  | - | - | - | 1.000 | 0.289 | 45367 | 127694 | 116887 |
|  | - | - | - | 1.100 | 0.318 | 49336 | 127326 | 113510 |
|  | - | - | - | 1.200 | 0.347 | 53212 | 126960 | 110233 |
|  | - | - | - | 1.300 | 0.376 | 56998 | 126595 | 107054 |
|  | - | - | - | 1.400 | 0.405 | 60695 | 126230 | 103970 |
|  | - | - | - | 1.500 | 0.434 | 64307 | 125867 | 100977 |
|  | - | - | - | 1.600 | 0.463 | 67835 | 125505 | 98074 |
|  | - | - | - | 1.800 | 0.492 | 71282 | 125144 | 95257 |
|  | - | - | - | 1.900 | 0.521 | 74649 | 124784 | 92524 |
|  | - | - | - | 2.000 | 0.579 | 77939 | 124425 | 89872 |
|  | - | - |  |  |  |  |  |  |
|  | - | -154 | 124067 | 87298 |  |  |  |  |

Catch and SSB are given in tonnes


Figure 3.1.1 Western Baltic Spring Spawning Herring. Catches and TACs by area. Top panel) Catches of Western Baltic Spring Spawning (WBSS) and North Sea Autumn Spawning (NSAS) herring in division IIIa, and the total TAC for both stocks. Middle panel) Catches and TACs of WBSS herring in subdivisions 22-24. Bottom panel) Total catch of WBSS herring in Div IVa, Div IIIa and SD 22-24.


Figure 3.6.1.1 Western Baltic Spring Spawning Herring. Proportion (by numbers) of a given age (in winter rings) in the catch.


Figure 3.6.1.2 Western Baltic Spring Spawning Herring. Proportion (by weight) of a given age (in winter rings) in the catch.


Figure 3.6.1.3 Western Baltic Spring Spawning Herring. Weight at age (in winter rings) in the stock.


Figure 3.6.1.4 Western Baltic Spring Spawning Herring. Time series of the individual index values used in the assessment, showing the German Acoustic survey (GerAS), the Herring acoustic survey (HerAS) and the N20 larval index.


Figure 3.6.4.1 Western Baltic Spring Spawning Herring. "Otolith" plot. The main figure depicts the uncertainty in the estimated spawning stock biomass and average fishing mortality, and their correlation. Contour lines give the $1 \%, 5 \%, 25 \%, 50 \%$ and $75 \%$ confidence intervals for the two estimated parameters and are estimated from a parametric bootstrap based on the variancecovariance matrix in the parameters returned by FLICA. The plots to the right and top of the main plot give the probability distribution in the SSB and mean fishing mortality respectively. The SSB and fishing mortality estimated by the method is plotted on all three plots with a heavy dot. $95 \%$ confidence intervals, with their corresponding values, are given on the plots to the right and top of the main plot.


Figure 3.6.4.2 Western Baltic Spring Spawning Herring. Stock summary plot. Top panel: Spawning stock biomass. Second panel: Recruitment (at age 0 -wr) as a function of time. Bottom panel:: Mean annual fishing mortality on ages 3-6 ringers as a function of time.

## Fitted catch diagnostics



Figure 3.6.4.3 Western Baltic Spring Spawning Herring. Diagnostics of selection pattern. a) Bubbles plot of log catch residuals by age (weighting applied) and year. Grey bubbles correspond to negative log residuals. The largest residual is given. b) Estimated selection parameters (relative to $\mathbf{4} \mathbf{w r}$ ) with $\mathbf{9 5 \%}$ confidence intervals. c): Marginal totals of residuals by year. d). Marginal totals of residuals by age (wr).


Figure 3.6.4.4 Western Baltic Spring Spawning Herring. Diagnostics of the German acoustic survey in subdivision 21-24 ("Ger AS 1-3 wr") fit at 1 wr from the assessment. a) Comparison of observed (points) and fitted (line) index value. b) Scatterplot of index observations versus FLICA estimates of stock numbers at age. Fitted catchability (linear model - solid line), with $95 \%$ confidence interval (dotted line). c) Log residuals of catchability model fitted by FLICA as a function of time. d). Log residuals from the catchability model against stock size at age estimated by the FLICA assessment method. e). Normal Q-Q plot of log residuals (points) with fitted linear regression (solid line) and $\mathbf{9 0} \%$ confidence interval for predication (dotted line).


Figure 3.6.4.5 Western Baltic Spring Spawning Herring. Diagnostics of the German acoustic survey in subdivision 21-24 ("Ger AS 1-3 wr") fit at 2 wr from the assessment. a) Comparison of observed (points) and fitted (line) index value. b) Scatterplot of index observations versus FLICA estimates of stock numbers at age. Fitted catchability (linear model - solid line), with $95 \%$ confidence interval (dotted line). c) Log residuals of catchability model fitted by FLICA as a function of time. d). Log residuals from the catchability model against stock size at age estimated by the FLICA assessment method. e). Normal Q-Q plot of log residuals (points) with fitted linear regression (solid line) and $\mathbf{9 0} \%$ confidence interval for predication (dotted line).


Figure 3.6.4.6 Western Baltic Spring Spawning Herring. Diagnostics of the German acoustic survey in subdivision 21-24 ("Ger AS 1-3 wr") fit at 3 wr from the assessment. a) Comparison of observed (points) and fitted (line) index value. b) Scatterplot of index observations versus FLICA estimates of stock numbers at age. Fitted catchability (linear model - solid line), with $95 \%$ confidence interval (dotted line). c) Log residuals of catchability model fitted by FLICA as a function of time. d). Log residuals from the catchability model against stock size at age estimated by the FLICA assessment method. e). Normal Q-Q plot of log residuals (points) with fitted linear regression (solid line) and $90 \%$ confidence interval for predication (dotted line).


Figure 3.6.4.7 Western Baltic Spring Spawning Herring. Diagnostics of the Herring acoustic survey in the North Sea and division IIIa ("HerAS 3-6 wr") fit at 3 wr from the assessment. a) Comparison of observed (points) and fitted (line) index value. b) Scatterplot of index observations versus FLICA estimates of stock numbers at age. Fitted catchability (linear model - solid line), with $\mathbf{9 5 \%}$ confidence interval (dotted line). c) Log residuals of catchability model fitted by FLICA as a function of time. d). Log residuals from the catchability model against stock size at age estimated by the FLICA assessment method. e). Normal Q-Q plot of log residuals (points) with fitted linear regression (solid line) and $90 \%$ confidence interval for predication (dotted line).


Figure 3.6.4.8 Western Baltic Spring Spawning Herring. Diagnostics of the Herring acoustic survey in the North Sea and division IIIa ("HerAS 3-6 wr") fit at 4 wr from the assessment. a) Comparison of observed (points) and fitted (line) index value. b) Scatterplot of index observations versus FLICA estimates of stock numbers at age. Fitted catchability (linear model - solid line), with $\mathbf{9 5 \%}$ confidence interval (dotted line). c) Log residuals of catchability model fitted by FLICA as a function of time. d). Log residuals from the catchability model against stock size at age estimated by the FLICA assessment method. e). Normal Q-Q plot of log residuals (points) with fitted linear regression (solid line) and $90 \%$ confidence interval for predication (dotted line).


Figure 3.6.4.9 Western Baltic Spring Spawning Herring. Diagnostics of the Herring acoustic survey in the North Sea and division IIIa ("HerAS 3-6 wr") fit at 5 wr from the assessment. a) Comparison of observed (points) and fitted (line) index value. b) Scatterplot of index observations versus FLICA estimates of stock numbers at age. Fitted catchability (linear model - solid line), with $\mathbf{9 5 \%}$ confidence interval (dotted line). c) Log residuals of catchability model fitted by FLICA as a function of time. d). Log residuals from the catchability model against stock size at age estimated by the FLICA assessment method. e). Normal Q-Q plot of log residuals (points) with fitted linear regression (solid line) and $\mathbf{9 0} \%$ confidence interval for predication (dotted line).


Figure 3.6.4.10 Western Baltic Spring Spawning Herring. Diagnostics of the Herring acoustic survey in the North Sea and division IIIa ("HerAS 3-6 wr") fit at 6 wr from the assessment. a) Comparison of observed (points) and fitted (line) index value. b) Scatterplot of index observations versus FLICA estimates of stock numbers at age. Fitted catchability (linear model - solid line), with $\mathbf{9 5 \%}$ confidence interval (dotted line). c) Log residuals of catchability model fitted by FLICA as a function of time. d). Log residuals from the catchability model against stock size at age estimated by the FLICA assessment method. e). Normal Q-Q plot of log residuals (points) with fitted linear regression (solid line) and $90 \%$ confidence interval for predication (dotted line).

## N20, age 0, diagnostics



Figure 3.6.4.11 Western Baltic Spring Spawning Herring. Diagnostics of the N20 larval index from the assessment. a) Comparison of observed (points) and fitted (line) index value. b) Scatterplot of index observations versus FLICA estimates of stock numbers at age. Fitted catchability (linear model - solid line), with $95 \%$ confidence interval (dotted line). c) Log residuals of catchability model fitted by FLICA as a function of time. d). Log residuals from the catchability model against stock size at age estimated by the FLICA assessment method. e). Normal Q-Q plot of log residuals (points) with fitted linear regression (solid line) and $\mathbf{9 0 \%}$ confidence interval for predication (dotted line).


Figure 3.6.4.12 Western Baltic Spring Spawning Herring. Mean contribution of a data point individual information groups (ages in each survey) to the FLICA objective function. The contribution is calculated from the mean of the squared residuals in the corresponding class, and weighted according to the appropriate value employed by the optimiser.


Figure 3.6.4.13 Western Baltic Spring Spawning Herring. Bubble plot showing the weighted residuals for each piece of fitted information. Individual values are weighted following the procedures employed internally with FLICA in calculating the objective function. The bubble scale is consistent between all panels.


Figure 3.6.4.14 Western Baltic Spring Spawning Herring. Analytical retrospective pattern in the assessment. Top panel: Spawning stock biomass. Middle panel: Recruitment at age 0 wr. Bottom panel: Mean fishing mortality in the ages 3-6 ringer. The heavy black line shows the current assessment.



Figure 3.6.4.16 Western Baltic Spring Spawning Herring. Stock-recruitment relationship. Recruitment at age $0-\mathrm{wr}$ (in thousands) is plotted as a function of spawning stock biomass (tonnes) estimated by the assessment. Successive years are joined by the line. Individual data points are labelled with the two-digit year.


Figure 3.6.5.1 Western Baltic Spring Spawning Herring. Contribution of each cohort (indicated by the colouring scheme, and the key to the right) to the spawning stock biomass.


Figure 3.6.5.2 Western Baltic Spring Spawning Herring. Relative contribution by weight of each cohort (indicated by the colouring scheme, and the key to the right) to the spawning stock biomass.

## 4 Celtic Sea and Division VIIj Herring

Exploratory Assessment with a final assessment presented
The assessment year for this stock runs from the $1^{\text {st }}$ April - $31^{\text {st }}$ March. Unless otherwise stated, year and year class are referred to by the first year in the season i.e. 2008 refers to the 2008/2009 season.

### 4.1 The Fishery

### 4.1.1 Advice and management applicable to 2008-2009

The TAC is set by calendar year and in 2008 was 7890 t , and in 2009 is 5917 t . In 2008, ICES considered the current stock size was uncertain but was likely to be as low as when the stock collapsed in the 1970s. At those recent levels of SSB there was a risk of reduced recruitment. Currently F was uncertain but too high and needed to be reduced. ICES recommended a rebuilding plan be put in place that would reduce catches. If no rebuilding plan was established, there should be no fishing. The rebuilding plan should be evaluated with respect to the precautionary approach.

## Rebuilding Plan

The Irish local fishery management committee developed a rebuilding plan for this stock. The Irish authorities submitted the plan to the European Commission in late 2008. STECF (2008) evaluated the plan, noting that it was likely to achieve its aims. The European Commission endorsed the plan, after it being noted in the TAC and quota regulation for 2009. The plan has been submitted to the Pelagic RAC. The RAC is working with the European Commission to develop a formal request for its evaluation by both ICES and STECF.
The plan (cited below) incorporated scientific advice with the main elements of the EU policy statement.

1 ) For 2009 , the TAC shall be reduced by $25 \%$ relative to the current year (2008).

2 ) In 2010 and subsequent years, the TAC shall be set equal to a fishing mortality of $\mathrm{F}_{0.1}$.
3 ) If, in the opinion of ICES and STECF, the catch should be reduced to the lowest possible level, the TAC for the following year will be reduced by $25 \%$.

4 ) Division VIIaS will be closed to herring fishing for 2009, 2010 and 2011.
5 ) A small-scale sentinel fishery will be permitted in the closed area, Division VIIaS. This fishery shall be confined to vessels, of no more than 65 feet in length. A maximum catch limitation of $8 \%$ of the Irish quota shall be exclusively allocated to this sentinel fishery.
6 ) Every three years from the date of entry into force of this Regulation, the Commission shall request ICES and STECF to evaluate the progress of this rebuilding plan.
7 ) When the SSB is deemed to have recovered to a size equal to or greater than $B_{p a}$ in three consecutive years, the rebuilding plan will be superseded by a long-term management plan.

### 4.1.2 The fishery in 2008/2009

In 2008-2009, 26 vessels took part in the Irish fishery. These are categorised as follows:

- 4 Pelagic refrigerated seawater (RSW) trawlers
- 6 Polyvalent bulk storage trawlers,
- 16 Polyvalent dry hold trawlers.

The fishery took place in the third quarter only in VIIj and in the fourth and first quarters in all three areas. Most vessels under 20 m reported landings of less than 100 $t$ for the season while a number of RSW vessels reported combined landings greater than 1100 t . The term "Polyvalent" refers to a segment of the Irish fleet, entitled to fish for any species to catch a variety of species, under Irish law.

The third quarter fishery took place in one statistical rectangle (31E0) in VIIj, landing a total of 360 t . The fourth quarter fishery began around the $1^{\text {st }}$ October, and lasted until the $2^{\text {nd }}$ week of December. The quarter 4 fishery took place in VIIj, off the south Irish coast, and further east in VIIg and between Cork and Capel Island and also further east in VIIaS.

As part of the rebuilding measures, the fishery was closed in quarter 1, 2009, except for $270 t$ allocation for a sentinel fishery.

The distribution of the total landings are presented in Figure 4.1.2.1

### 4.1.3 The catches in 2008/2009

The estimated national catches from 1988-2008 for the combined areas by year and by season ( $1^{\text {st }}$ April-31 ${ }^{\text {st }}$ March) are given in Table 4.1.3.1 and Table 4.1.3.2 respectively. The catch, taken during the 2008 season has fallen to the lowest estimate in the entire series, about 5800 t (Figure 4.1.3.1.). The catch data include discards, until 1997. Official catches reported from other nations are assumed to be taken from other areas and are subtracted as unallocated catches.

There are no recent estimates of discards for this fishery. Anecdotal reports from fishermen suggest that discarding is not a feature of this fishery at present.

### 4.1.4 Regulations and their effects

The closure of VIIaS, except for a sentinel fishery means that only small dry hold vessels, no more than 65 feet total length, can fish in that area. This closure has meant that the majority of the quota was taken by the larger bulk storage vessels further west, including VIIj. There have been two closures of VIIaS (2002-2003 and 2007present). Though it is difficult to assess their effectiveness it can be seen that in each period of closure, F has been substantially reduced and SSB grew (Section 4.6). This area, particularly the area off Dunmore East, is important for recruit spawners. It can be expected that the closure allows these fish to spawn at least once, and contribute to SSB through further growth and spawning potential.

### 4.1.5 Changes in fishing technology and fishing patterns

The stock is exploited by three types of vessels, larger boats with RSW or bulk storage and smaller dry hold vessels. The smaller vessels are confined to the spawning grounds (VIIaS and VIIg) during the winter period. The refrigerated seawater (RSW) tank vessels target the stock inshore in winter and offshore during the summer feed-
ing phase (VIIg), but are not allowed in VIIaS under the terms of the rebuilding plan. In the past season there has been increased fishing VIIj.

The sentinel fishery that was allowed in VIIaS closed area mainly took place in Waterford harbour mouth and roads, rather than upstream.

### 4.2 Biological composition of the catch

### 4.2.1 Catches in numbers-at-age

Catch numbers at age are available for the period 1958 to 2008. In 2008, there was a strong dominance of 2 ringers (2005 year class) and 4-ringers (2003 year class). This cohort was strong in the previous season as 3-ringers also. The weak 2001/2002 year class has almost disappeared from the catches by now (Table 4.2.1.1). The catch numbers at age for ages 1-9 are presented in Figure 4.2.1.1 and it can be seen that there has been a truncation of older ages in recent years. Due to this truncation a new plus group is now set at $6+$ and the yearly mean standardised plot is shown in Figure 4.2.1.2. Both plots show that 2-ringers have been the dominant age in catches in general throughout the series .

The overall proportions at age were similar in all sampled metiers (division*quarter). However, unusually the survey and the commercial fishery did not agree well in terms of proportions at age (Figure 4.2.1.3). The 4 -ringers that were so dominant in the commercial catch were less dominant in the survey, and 0-ringers were found in the latter only. Apart from these two age groups, the patterns are similar. The 2003 year class appeared stronger in the fishery that was mainly inshore during the time of the survey.

Unlike recent years, samples from VIIaS within Waterford Harbour and outside were not raised separately. This is because most of the fishing took place outside, on the Roads of Waterford.

Table 4.2.1.2 shows the length frequency data by area and quarter. A similar length range was found in each area.

### 4.2.2 Quality of catch and biological data

Biological sampling of the catches throughout the region was comprehensive throughout the area exploited by the Irish fishery (Table 4.2.2.1). However no samples accompanied reported landings from the Netherlands, Germany or France. Under the Data Collection Programme the sampling of this stock is well above that required by the Minimum Programme (Section 1.5).

The quality of catch data has varied over time. A rudimentary history of the Irish fishery since 1958 is presented in the Stock Annex. The quality of the landings data has improved in most recent years, particularly since 2004, when a low tolerance for water in catches was introduced. In 2008/2009 only preliminary data were available at the time of the working group. Best estimates of small boat catches were used for the VIIaS sentinel fishery. This is because not all the vessels are required to make logbook returns, being less than 10 m in total length.

Discarding was a major feature of the fishery from 1983 to 1997, when the fishery sought fish of a particular roe quality, discarding early stage, spent and young fish. Though discarding (slippage) is thought to be lower in subsequent years, the tight quota situation coupled with market requirements are known to lead to some dis-
carding, particularly of smaller fish. There is no information on misreporting in this fishery in recent years, but it is thought to have decreased.

### 4.3 Fishery Independent Information

### 4.3.1 Acoustic Surveys

Since 2005 this survey has had a standard design and is conducted at the same time each year. The surveys carried out from 2002 and 2003 are comparable with the later surveys. This series dates from 1995 and is presented in Table 4.3.1.1.

The acoustic survey of the 2008/2009 season was carried out in October 2008, on the Celtic Explorer (O'Donnell, et al 2008). The survey track began at the northern boundary of VIIj, covering the SW bays in zig-zags and parallel transects (Figure 4.3.1.1a). As in previous seasons, very little herring was registered in the bays of VIIj Figure (4.3.1.1b). The main broad scale survey in VIIg and VIIaS adapted a parallel transect design and showed the greatest concentrations of herring.

In 2008/2009 the SSB estimate was 90855 t . This is an increase of about $96 \%$, from the previous year. The current estimate is also much more precise, with a CV of $20 \%$, the most precise in the series. This estimate is associated with more even distribution of herring than in previous years.

### 4.3.2 Other surveys

In 2008, a pair trawl survey was conducted, to find juvenile herring (Clarke et al. 2008). This was a scoping exercise, to map the distribution of juvenile herring in advance of developing a recruit index for this stock. A new Irish recruit survey is envisaged by the industry initiated plan to rebuild the stock. It is envisaged that a combination of the Northern Ireland GFS and Irish survey of the Celtic Sea and VIIj could be used. GFS surveys could provide useful indices for the component of the stock, if the origin of herring in catches can be identified. Some progress has been made on this (Beggs, 2008 WD).

### 4.4 Mean weights-at-age and maturity-at-age

The mean weights in the catch and mean weight in the stock at spawning time are presented in Figure 4.4.1.1 and 4.4.1.2 respectively. There has been an overall downward trend in mean weights at age since the mid-1980s. However, the values for 2008/2009 have shown an increase. The 2-ringers have shown an increase since 2005.

Mean weights in the stock at spawning time were calculated from biological samples, for quarters 4 and 1 (Figure 4.4.1.2). A slight increase is evident in these data for the most recent season.

The 1 ringers that are resident in the Celtic Sea appear to have greater than $50 \%$ maturity. The Celtic Sea 1 ringers that are present in the Irish Sea have less than $50 \%$ maturity (Beggs WD, 2009).

### 4.5 Recruitment

At present there are no recruitment estimates for this stock that can be used for predictive purposes. The issue of mixing of Irish Sea and Celtic Sea juveniles means that recruitment is not well represented in the catch data.

### 4.6 Assessment

### 4.6.1 Exploratory Assessments

This stock was benchmarked in 2007. Problems with conflicting signals in input data and changes in the fishing pattern prevented a final assessment being conducted. The assessment was considered indicative of trends.

A number of further exploratory runs were conducted in 2009 using FLICA. The details of these runs are presented in Table 4.6.1.1. The main settings examined were as follows

- Plus group reduced to $6+$ from $7+$
- Shortening of the survey time series from 1995-2008 to 2002-2008
- Terminal selection of 1, 1.1, and 1.2.
- Changing the separable period from 6 years to 5 years and 4 years

The catch and survey residual patterns are shown from the 6+ and 7+ runs in Figure 4.6.1.1 and catch diagnostics and selection patterns from each run are presented in Figure 4.6.1.2. The $6+$ run shows a better residual pattern, with smaller residuals, and a flat topped selection pattern.

The length of the survey time series was also examined. When the full series from 1995 was used we can see a much improved residual pattern displayed for 2002 onwards (Figure 4.6.1.3). From 2005 a uniform track design has been adopted and the survey timing is fixed in October each year. The residuals for the 2000 and 2001 surveys are the biggest and it was decided to test the removal of these noisy data and use the survey series from 2002 - 2008 in further runs. Some year effects are evident in the shorter time series but the residuals are small overall.

Changing the selection pattern using values of 1, 1.1 and 1.2 for the terminal age was examined and it was found that increasing the selection on the oldest age did not show a significant improvement. For further runs 1 was used because there is no evidence from the fishery that selection is higher for older ages. The precision of the assessment is reduced when the selection on the oldest age is increased.

The reduction of the separable period to 4 and 5 years from 6 years probably led to over parameterisation of the model, and also resulted in an inflated stock size and very low F values. A more precise estimate of SSB is achieved when a longer separable period is used. The SSB confidence limits are narrower using a longer separable period (Figure 4.6.1.4). The separable period of 6 years was used for all other runs.

The estimates of Mean F and SSB for some of the exploratory runs are presented in Figure 4.6.1.4. The perception of the stock is similar for all runs except when the separable period is shortened. The otolith plots produced form the $6+$ and $7+$ run are presented in Figure 4.6.1.5 and show very similar values. The exploratory runs have shown that using the $6+$ data and the survey time series from 2002-2008 produces improved residual patterns.

Historical retrospective assessments are presented in Figure 4.6.1.6. These are based on final "SPALY" ICA exploratory assessments from recent years, with terminal year SSB adjusted to account for poor 1-ringer estimation (see Stock Annex). These show a relatively balanced pattern, with no systematic bias in estimation of stock parameters in recent years. In the historical period there are retrospective patterns, with the 2009 accepted assessment producing a lower ( 24000 t ) estimate of $\mathrm{B}_{\text {loss. }}$. This new estimate
of Bloss is less than $8 \%$ lower than $B_{\lim }$ (Section 4.9) and is due to the reduction of the plus group to $6+$.
The historical retrospectives were performed using the standard procedure for this stock, where the terminal SSB is adjusted by using GM recruitment. The analytical retrospectives and precision estimates (Figures 4.6.1.5 and 4.6.2.5) are not adjusted with GM recruitment. However they were considered to be good indicators of the best ICA run. The decision regarding the best assessment run was based on model fit diagnostics, and not on stock trajectories.
The results from the exploratory assessment show that there is improved consistency and precision in this assessment. The working group has decided to accept this assessment as an analytical assessment. This accepted assessment is largely based on the same settings as those from the previous benchmark. At the time of the benchmark, the series of comparable surveys was too short and the separable period included a marked change in fishing pattern (2003). In 2009, these problems were resolved. The separable period now does not include the changing pattern. Several years of comparable surveys (2002-present) are now available for tuning.

### 4.6.2 Final assessment

Based on the explorations carried out above a final run was chosen which used plus group at $6+$. The input and output data are presented in Table 4.6.2.1 to 4.6.2.1. The survey series from 2002-2008 was used as well as a 6 year separable period with terminal selection set at 1 relative to 3 ring. The diagnostics are presented in Figure 4.6.2.1 to Figure 4.6.2.6.

The analytical retrospective pattern is displayed in Figure 4.6.2.5. The retrospective was fitted as far back as 2003 but excludes the 2004 estimates. A retrospective analysis cannot be extended into earlier years because of the lack of reliable survey data. However it can be seen from the years presented that there is no systematic bias and there is a reasonably tight pattern of evenly balanced retrospective estimations.

### 4.6.3 State of the stock

The stock appears to have increased in size and is above $B_{\text {pa. }}$ F has declined from the peak in 2003, and is estimated to be below $\mathrm{F}_{0.1}$. Overall recruitment is around long term mean. The stock is showing signs of recovery. However it is still very dependent on strength of incoming year classes, that cannot be observed until fully recruited.

### 4.7 Short term projections

### 4.7.1 Deterministic Short Term Projections

A deterministic short term forecast was performed, using the MFDP software (Smith, 2000). The input data are presented in Table 4.7.1.1. Geometric mean (1995-2006) was used because this represents a period where recruitment has been fluctuating around the mean. Mean weights in the catch and in the stock were calculated as means over the last three years. Recruits (1-ring) are poorly represented in the catch and only one observation of their abundance is available. Yet $50 \%$ of these are considered mature and they make an important contribution to the SSB. The population numbers at 1 ring are replaced by geometric mean. Population numbers of 2 ringers in the intermediate season (2009) were calculated by the degradation of geometric mean recruitment (1995-2006) using the equation below.
$N_{t+1}=N_{t}^{*} e^{-\mathrm{Ft}+\mathrm{Mt}}$
The short term forecast was performed using the predicted catch in the interim year 2009. This was calculated as the remaining Irish quota for 2009 + the likely Irish catch in quarter 1 of 2010. The 2010 quarter 1 catch was estimated assuming that the TAC would be increased by $15 \%$ using the EU TAC Decision Rule for stocks for which STECF advice is that the stock is increasing. The use of Irish catch estimates in the interim year assumes that other countries' catches are unallocated.

The results of the short term projection are presented in Table 4.7.1.2 and Table 4.7.1.3. Fishing according to the proposed rebuilding plan implies catches of $9,227 \mathrm{t}$ in 2010. Only very high catches are associated with SSB < $\mathrm{B}_{\mathrm{pa}}$ in 2010.

### 4.7.2 Yield Per Recruit

A yield per recruit analysis was conducted using MFYPR. The yield per recruit curve is presented in Figure 4.7.2.1 and $\mathrm{F}_{0.1}$ was estimated to be 0.17.

### 4.8 Medium term projections

No medium term projections were conducted by the working group.

### 4.9 Precautionary and yield based reference points

Reference points are defined for this stock, $\mathbf{B}_{\mathrm{pa}}$ is currently at 44000 t (low probability of low recruitment) and $\mathbf{B}_{\text {lim }}$ at 26000 t ( $\mathbf{B}_{\text {loss }}$ ) for this stock. $\mathbf{F}_{\text {pa }}$ and $\mathbf{F}_{\text {lim }}$ are not defined. $\mathrm{F}_{\mathrm{msy}}$ has not been estimated. However $\mathrm{F}_{0.1}$ can be assumed to be a proxy for $\mathrm{F}_{\text {msy }}$ and was estimated in 2009 to be $=0.17$ (Section 4.8).

The historical retrospective analysis (Section 4.6.1) shows that the accepted assessment produces a downward revision ( $<8 \%$ ) of the basis for Blim, which is Bloss. This is due to the reduction in the plus group. It is not considered necessary to carry out a downward revision of reference points because the revision is small. Also there is some evidence that Blim should be revised upwards, to the point of recruitment impairment estimated by Clarke and Egan (2008). These authors showed a changepoint in a segmented regression at 47000 t .

### 4.10 Quality of the Assessment

A final analytical assessment is being proposed. The precision of the assessment estimated through a parametric bootstrap routine are presented for the $6+$ assessment and also the 7+ assessment in Figure 4.6.1.5. Both display similar ranges for SSB and mean F for 2-5 ring.

There is improved coherence between the catch at age and the survey data. The survey results are more stable since 2002. Since 2005 a uniform design was adopted. This improved coherence is reflected in better tuning diagnostics.

### 4.11 Management Considerations

Fishing mortality on this stock was high for many years, well above a long term sustainable level of $\mathrm{F}_{0.1}=0.17$. In the past three years F has been substantially reduced and is now below $\mathrm{F}_{0.1}$ and at its lowest rate in 45 years. This is associated with reduced catches and management actions which led to the closure of the fishery in quarter 1, 2009. In the past two years, ICES has recommended that catches of around $5,000 \mathrm{t}$
would be associated with stock recovery. The current landings estimate ( $5,700 \mathrm{t}$ ) is close to that suggested catch and is the lowest in the series.

The WG is aware of the rebuilding plan proposed by the Irish industry. This envisaged a $25 \%$ TAC reduction in 2009, with future catch levels based on an $\mathrm{F}_{0.1}$ strategy. The short term forecast conducted by the working group shows that fishing at $\mathrm{F}_{0.1}$ would imply catches of over 9,000 t in 2010. It is unclear if management will follow the rebuilding plan. However there is good evidence to show that the stock has increased substantially. Consequently, the current advice for zero catch is no longer justified. The rebuilding plan should continue until 2011 and then if the stock can be shown to have rebuilt, the rebuilding plan will be replaced by a long term management plan.

The measures to protect first time spawners by closing the VIIaS Box should continue until 2011 as set out in the rebuilding plan. The measure has not been in place long enough to assess its benefits fully. Sampling of the sentinel fishery which takes place in this closed area will continue.

### 4.12 Environment

## Ecosystem considerations

Herring are an important prey species in the ecosystem and also one of the dominant planktivorous fish.

The spawning grounds for herring in the Celtic Sea are well known and are located inshore close to the coast. These spawning grounds may contain one or more spawning beds on which herring deposit their eggs. Individual spawning beds within the spawning grounds have been mapped and consist of either gravel or flat stone (Breslin, 1998). Spawning grounds tend to be vulnerable to anthropogenic influences such as dredging, sand and gravel extraction, dumping of dredge spoil and waste from fish cages. There have been several proposals for extraction of gravel and to dump dredge spoil in recent years. Many of these proposals relate to known herring spawning grounds. ICES has consistently advised that activities that perturb herring spawning grounds should be avoided.

Herring fisheries tend to be clean with little bycatch of other fish. Mega fauna by catch is unquantified, though anecdotal reports suggest that seals are caught from time to time.

## Changes in the environment

Temperatures in this area have been increasing over the last number of decades. There are indications that salinity is also increasing (ICES 2006). It is considered that this could have implications for herring, that is at the southern edge of its distribution in this area. It is known that similar environmental changes have affected the North Sea herring. There is no evidence that changes in the environmental regime in the Celtic Sea has had any effect on productivity of this stock.

Table 4.1.3.1. Celtic Sea and Division VIIj herring. Landings by quota year (t), 1988-2008. (Data provided by Working Group members.) These figures may not in all cases correspond to the official statistics and cannot be used for management purposes.

| Year | France | Germany | Ireland | Netherlands | U.K. | Unallocated | Discards | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | - | - | 16,800 | - | - | - | 2,400 | 19,200 |
| 1989 | + | - | 16,000 | 1,900 | - | 1,300 | 3,500 | 22,700 |
| 1990 | + | - | 15,800 | 1,000 | 200 | 700 | 2,500 | 20,200 |
| 1991 | + | 100 | 19,400 | 1,600 | - | 600 | 1,900 | 23,600 |
| 1992 | 500 | - | 18,000 | 100 | + | 2,300 | 2,100 | 23,000 |
| 1993 | - | - | 19,000 | 1,300 | + | -1,100 | 1,900 | 21,100 |
| 1994 | + | 200 | 17,400 | 1,300 | + | -1,500 | 1,700 | 19,100 |
| 1995 | 200 | 200 | 18,000 | 100 | + | -200 | 700 | 19,000 |
| 1996 | 1,000 | 0 | 18,600 | 1,000 | - | -1,800 | 3,000 | 21,800 |
| 1997 | 1,300 | 0 | 18,000 | 1,400 | - | -2,600 | 700 | 18,800 |
| 1998 | + | - | 19,300 | 1,200 | - | -200 | - | 20,300 |
| 1999 |  | 200 | 17,900 | 1300 | + | -1300 | - | 18,100 |
| 2000 | 573 | 228 | 18,038 | 44 | 1 | -617 | - | 18,267 |
| 2001 | 1,359 | 219 | 17,729 | - | - | -1578 | - | 17,729 |
| 2002 | 734 | - | 10,550 | 257 | - | -991 | - | 10,550 |
| 2003 | 800 | - | 10,875 | 692 | 14 | -1,506 | - | 10,875 |
| 2004 | 801 | 41 | 11,024 | - | - | -801 | - | 11,065 |
| 2005 | 821 | 150 | 8452 | 799 | - | -1770 | - | 8,452 |
| 2006 | - | - | 8,530 | 518 | 5 | -523 | - | 8,530 |
| 2007 | 581 | 248 | 8,268 | 463 | 63 | -1355 | - | 8,268 |
| 2008 | 503 | 191 | 6,774 | 291 |  | -985 | - | 6,774 |

Table 4.1.3.2. Celtic Sea \& Division VIIj herring landings ( $t$ ) by assessment year (1st April-31st March) 1988/1989-2008/2009. (Data provided by Working Group members.) These figures may not in all cases correspond to the official statistics and cannot be used for management purposes.

| Year | France | Germany | Ireland | Netherlands | U.K. | Unallocated | Discards | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988/1989 | - | - | 17,000 | - | - | - | 3,400 | 20,400 |
| 1989/1990 | + | - | 15,000 | 1,900 | - | 2,600 | 3,600 | 23,100 |
| 1990/1991 | + | - | 15,000 | 1,000 | 200 | 700 | 1,700 | 18,600 |
| 1991/1992 | 500 | 100 | 21,400 | 1,600 | - | -100 | 2,100 | 25,600 |
| 1992/1993 | - | - | 18,000 | 1,300 | - | -100 | 2,000 | 21,200 |
| 1993/1994 | - | - | 16,600 | 1,300 | + | -1,100 | 1,800 | 18,600 |
| 1994/1995 | + | 200 | 17,400 | 1,300 | + | -1,500 | 1,900 | 19,300 |
| 1995/1996 | 200 | 200 | 20,000 | 100 | + | -200 | 3,000 | 23,300 |
| 1996/1997 | 1,000 | - | 17,900 | 1,000 | - | -1,800 | 750 | 18,800 |
| 1997/1998 | 1,300 | - | 19,900 | 1,400 | - | -2100 | - | 20,500 |
| 1998/1999 | + | - | 17,700 | 1,200 | - | -700 | - | 18,200 |
| 1999/2000 |  | 200 | 18,300 | 1300 | + | -1300 | - | 18,500 |
| 2000/2001 | 573 | 228 | 16,962 | 44 | 1 | -617 | - | 17,191 |
| 2001/2002 | - | - | 15,236 | - | - | - | - | 15,236 |
| 2002/2003 | 734 | - | 7,465 | 257 | - | -991 | - | 7,465 |
| 2003/2004 | 800 | - | 11,536 | 610 | 14 | -1,424 | - | 11,536 |
| 2004/2005 | 801 | 41 | 12,702 | - | - | -801 | - | 12,743 |
| 2005/2006 | 821 | 150 | 9,494 | 799 | - | -1770 | - | 9,494 |
| 2006/2007 | - | - | 6,944 | 518 | 5 | -523 | - | 6,944 |
| 2007/2008 | 379 | 248 | 7,636 | 327 | - | -954 | - | 7,636 |
| 2008/2009 | 503 | 191 | 5,793 | 150 |  | -844 | - | 5,793 |

Table 4.2.1.1. Celtic Sea \& Division VIIj herring. Comparison of age distributions (percentages) in the catches of Celtic Sea and VIIj herring over the time series.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1958 | 1\% | 3\% | 25\% | 20\% | 10\% | 18\% | 12\% | 7\% | 4\% |
| 1959 | 1\% | 27\% | 2\% | 20\% | 12\% | 6\% | 19\% | 4\% | 8\% |
| 1960 | 2\% | 53\% | 18\% | 3\% | 10\% | 3\% | 4\% | 3\% | 3\% |
| 1961 | 3\% | 22\% | 44\% | 8\% | 3\% | 7\% | 4\% | 2\% | 7\% |
| 1962 | 1\% | 16\% | 17\% | 41\% | 7\% | 3\% | 7\% | 3\% | 5\% |
| 1963 | 0\% | 52\% | 13\% | 4\% | 21\% | 3\% | 1\% | 3\% | 3\% |
| 1964 | 12\% | 25\% | 28\% | 11\% | 3\% | 14\% | 2\% | 1\% | 4\% |
| 1965 | 0\% | 56\% | 8\% | 13\% | 3\% | 4\% | 10\% | 1\% | 6\% |
| 1966 | 5\% | 15\% | 46\% | 8\% | 10\% | 4\% | 3\% | 7\% | 3\% |
| 1967 | 5\% | 26\% | 13\% | 32\% | 6\% | 6\% | 3\% | 4\% | 4\% |
| 1968 | 8\% | 35\% | 25\% | 7\% | 14\% | 3\% | 3\% | 1\% | 3\% |
| 1969 | 4\% | 40\% | 24\% | 14\% | 5\% | 8\% | 2\% | 1\% | 1\% |
| 1970 | 1\% | 24\% | 33\% | 17\% | 12\% | 5\% | 4\% | 1\% | 2\% |
| 1971 | 8\% | 15\% | 24\% | 27\% | 12\% | 7\% | 3\% | 3\% | 1\% |
| 1972 | 4\% | 67\% | 9\% | 8\% | 7\% | 2\% | 1\% | 1\% | 0\% |
| 1973 | 16\% | 26\% | 38\% | 5\% | 7\% | 4\% | 2\% | 2\% | 1\% |
| 1974 | 5\% | 43\% | 17\% | 22\% | 4\% | 4\% | 3\% | 1\% | 1\% |
| 1975 | 18\% | 22\% | 25\% | 11\% | 13\% | 5\% | 2\% | 2\% | 2\% |
| 1976 | 26\% | 22\% | 14\% | 14\% | 6\% | 9\% | 4\% | 2\% | 3\% |
| 1977 | 20\% | 31\% | 22\% | 13\% | 4\% | 5\% | 3\% | 1\% | 1\% |
| 1978 | 7\% | 35\% | 31\% | 14\% | 4\% | 4\% | 1\% | 2\% | 1\% |
| 1979 | 21\% | 26\% | 23\% | 16\% | 5\% | 2\% | 2\% | 1\% | 1\% |
| 1980 | 11\% | 47\% | 18\% | 10\% | 4\% | 3\% | 2\% | 2\% | 1\% |
| 1981 | 40\% | 22\% | 22\% | 6\% | 5\% | 4\% | 1\% | 0\% | 1\% |
| 1982 | 20\% | 55\% | 11\% | 6\% | 2\% | 2\% | 2\% | 0\% | 1\% |
| 1983 | 9\% | 68\% | 18\% | 2\% | 1\% | 0\% | 0\% | 1\% | 0\% |
| 1984 | 11\% | 53\% | 24\% | 9\% | 1\% | 1\% | 0\% | 0\% | 0\% |
| 1985 | 14\% | 44\% | 28\% | 12\% | 2\% | 0\% | 0\% | 0\% | 0\% |
| 1986 | 3\% | 39\% | 29\% | 22\% | 6\% | 1\% | 0\% | 0\% | 0\% |
| 1987 | 4\% | 42\% | 27\% | 15\% | 9\% | 2\% | 1\% | 0\% | 0\% |
| 1988 | 2\% | 61\% | 23\% | 7\% | 4\% | 2\% | 1\% | 0\% | 0\% |
| 1989 | 5\% | 27\% | 44\% | 13\% | 5\% | 2\% | 2\% | 0\% | 0\% |
| 1990 | 2\% | 35\% | 21\% | 30\% | 7\% | 3\% | 1\% | 1\% | 0\% |
| 1991 | 1\% | 40\% | 24\% | 11\% | 18\% | 3\% | 2\% | 1\% | 0\% |
| 1992 | 8\% | 19\% | 25\% | 20\% | 7\% | 13\% | 2\% | 5\% | 0\% |
| 1993 | 1\% | 72\% | 7\% | 8\% | 3\% | 2\% | 5\% | 1\% | 0\% |
| 1994 | 10\% | 29\% | 50\% | 3\% | 2\% | 4\% | 1\% | 1\% | 0\% |
| 1995 | 6\% | 49\% | 14\% | 23\% | 2\% | 2\% | 2\% | 1\% | 1\% |
| 1996 | 3\% | 46\% | 29\% | 6\% | 12\% | 2\% | 1\% | 1\% | 1\% |
| 1997 | 3\% | 26\% | 37\% | 22\% | 6\% | 4\% | 1\% | 1\% | 0\% |
| 1998 | 5\% | 34\% | 22\% | 23\% | 11\% | 3\% | 2\% | 0\% | 0\% |
| 1999 | 11\% | 27\% | 28\% | 11\% | 12\% | 7\% | 1\% | 2\% | 0\% |
| 2000 | 7\% | 58\% | 14\% | 9\% | 4\% | 5\% | 2\% | 0\% | 0\% |
| 2001 | 12\% | 49\% | 28\% | 5\% | 3\% | 1\% | 1\% | 0\% | 0\% |
| 2002 | 6\% | 46\% | 32\% | 9\% | 2\% | 2\% | 1\% | 0\% | 0\% |
| 2003 | 3\% | 41\% | 27\% | 16\% | 6\% | 4\% | 3\% | 0\% | 1\% |
| 2004 | 5\% | 10\% | 50\% | 24\% | 9\% | 2\% | 1\% | 0\% | 0\% |
| 2005 | 19\% | 38\% | 7\% | 23\% | 9\% | 2\% | 1\% | 0\% | 0\% |
| 2006 | 3\% | 58\% | 19\% | 4\% | 11\% | 4\% | 1\% | 0\% | 0\% |
| 2007 | 12\% | 17\% | 56\% | 9\% | 2\% | 3\% | 1\% | 0\% | 0\% |
| 2008 | 3\% | 31\% | 20\% | 38\% | 6\% | 1\% | 1\% | 0\% | 0\% |

Table 4.2.1.2. Celtic Sea \& Division VIIj herring. Length frequency distributions of the Irish catches (raised numbers in '000s) in the 2008/2009 season in the Celtic Sea and VIIj fishery.

|  | 2008 |  |  | 2009 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 7j Q3 | 7j Q4 | 7g Q4 | 7aS Q4 | 7j Q1 | 7g Q1 | 7aSQ1 | Total |
| 18 |  |  |  |  |  |  |  |  |
| 18.5 |  |  |  | 4 |  |  |  | 4 |
| 19 |  |  | 7 |  |  | 3 |  | 10 |
| 19.5 |  |  | 34 |  |  | 6 |  | 40 |
| 20 | 2 | 7 | 55 | 7 | 0 | 10 | 3 | 85 |
| 20.5 | 5 | 14 | 103 | 4 | 0 | 12 | 2 | 140 |
| 21 | 12 | 36 | 172 | 26 | 1 | 18 |  | 265 |
| 21.5 | 9 | 29 | 255 | 45 | 1 | 27 | 2 | 367 |
| 22 | 23 | 71 | 427 | 82 | 2 | 53 | 6 | 665 |
| 22.5 | 30 | 93 | 951 | 138 | 3 | 66 | 16 | 1297 |
| 23 | 80 | 243 | 1737 | 354 | 7 | 78 | 34 | 2531 |
| 23.5 | 108 | 328 | 2288 | 433 | 9 | 104 | 35 | 3305 |
| 24 | 152 | 464 | 2495 | 552 | 13 | 120 | 38 | 3835 |
| 24.5 | 246 | 749 | 3074 | 582 | 21 | 134 | 51 | 4857 |
| 25 | 309 | 942 | 3729 | 619 | 26 | 186 | 61 | 5872 |
| 25.5 | 400 | 1220 | 3880 | 634 | 34 | 171 | 61 | 6401 |
| 26 | 384 | 1170 | 2874 | 526 | 33 | 157 | 40 | 5184 |
| 26.5 | 269 | 821 | 1558 | 261 | 23 | 111 | 16 | 3058 |
| 27 | 140 | 428 | 669 | 108 | 12 | 88 | 10 | 1454 |
| 27.5 | 84 | 257 | 317 | 60 | 7 | 48 | 5 | 778 |
| 28 | 28 | 86 | 165 | 19 | 2 | 36 |  | 336 |
| 28.5 | 19 | 57 | 48 | 7 | 2 | 11 |  | 144 |
| 29 | 12 | 36 | 21 |  | 1 | 5 |  | 75 |

Table 4.2.2.1 Celtic Sea \& Division VIIj (2008/2009). Sampling intensity of Irish commercial catches. Only Ireland provides samples of this stock.

| ICES area | Year | Quarter | Landings (t) | No. Samples | No. aged | No. Measured | Aged/1000 t |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| VIIg | 2008 | 4 | 3473 | 16 | 1195 | 3607 | 344 |
| VIIg | 2009 | 1 | 188 | 12 | 293 | 3723 | 1558 |
|  |  |  |  |  |  |  |  |
| Sub-total |  |  | 3661 | 28 | 1488 | 7330 |  |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
| VIIaS | 2008 | 4 | 580 | 8 | 594 | 1196 | 1024 |
| VIIaS | 2009 | 1 | 46 | 2 | 149 | 237 | 3239 |
|  |  |  |  |  |  |  |  |
| Sub-total |  |  | 626 | 10 | 743 | 1433 |  |
|  |  |  |  |  |  |  |  |
| VIIj | 2008 | 3 | 364 | 0 | 516 | 988 | 1417 |
| VIIj | 2008 | 4 | 1111 | 7 | 516 | 988 | 464 |
| VIIj | 2009 | 1 | 31 | 0 | 516 | 988 | 16645 |
|  |  |  |  |  |  |  |  |
| Sub-total |  |  | 1506 | 7 | 1548 | 2964 |  |
|  |  |  |  |  |  |  |  |
| Total Celtic Sea |  |  | 5794 | 45 | 3779 | 11727 |  |

Table 4.3.1.1. Celtic Sea \& Division VIIj herring. Revised acoustic index of abundance. Total stock numbers-at-age ( $\mathbf{1 0}^{6}$ ) estimated using combined acoustic surveys (age refers in winter rings, biomass and SSB in 000's tonnes).

|  | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 0 | 202 | 3 | - | 0 | - | 25 | 40 | 0 | 24 | - | 2 | - | 1 | 99 |
| 1 | 25 | 164 | - | 30 | - | 102 | 28 | 42 | 13 | - | 65 | 21 | 106 | 64 |
| 2 | 157 | 795 | - | 186 | - | 112 | 187 | 185 | 62 | - | 137 | 211 | 70 | 295 |
| 3 | 38 | 262 | - | 133 | - | 13 | 213 | 151 | 60 | - | 28 | 48 | 220 | 111 |
| 4 | 34 | 53 | - | 165 | - | 2 | 42 | 30 | 17 | - | 54 | 14 | 31 | 162 |
| 5 | 5 | 43 | - | 87 | - | 1 | 47 | 7 | 5 | - | 22 | 11 | 9 | 27 |
| 6 | 3 | 1 | - | 25 | - | 0 | 33 | 7 | 1 | - | 5 | 1 | 13 | 6 |
| 7 | 1 | 15 | - | 24 | - | 0 | 24 | 3 | 0 | - | 1 | - | 4 | 5 |
| 8 | 2 | 0 | - | 4 | - | 0 | 15 | 0 | 0 | - | 0 | - | 1 |  |
| 9 | 2 | 2 | - | 2 | - | 0 | 52 | 0 | 0 | - | 0 | - | 0 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  | - |
| Abundance | 469 | 1338 | - | 656 |  | 256 | 681 | 423 | 183 | - | 312 | 305 | 454 | 769 |
| SSB | 36 | 151 |  | 100 |  | 20 | 95 | 41 | 20 | - | 33 | 36 | 46 | 90 |
| CV | 53 | 26 |  | 36 |  | 100 | 88 | 49 | 34 | - | 48 | 35 | 25 | 20 |
| Design | AR | AR |  | AR |  | AR | AR | AR | AR |  | R | R | R | R |

*AR Adaptive random; R random

Table 4.6.1.1. Celtic Sea and VIIj herring. Settings used in exploratory FLICA assessments.

|  |  |  | Plus |  |
| :--- | :--- | :--- | :--- | :--- |
|  | Survey Time Series | Selection | Group | Sep period |
| 1 | $2005-2008$ | 1 | 7 | 6 |
| 2 | $2002-2008$ | 1 | 7 | 6 |
| 3 | $1995-2008$ | 1.2 | 7 | 6 |
| 4 | $1995-2008$ | 1.1 | 7 | 6 |
| 5 | $1995-2008$ | 1 | 7 | 4 |
| 6 | $1995-2008$ | 1 | 7 | 5 |
| 7 | $1995-2008$ | 1 | 6 | 6 |
| 8 | $1995-2008$ | 1 | 6 | 6 |
| 9 | $2002-2008$ | 1 | 6 | 4 |
| 10 | $1995-2008$ | 1 | 6 | 5 |
| 11 | $1995-2008$ | 1 | 6 | 5 |
| 12 | $2002-2008$ | 1 | 6 | 4 |
| 13 | $2002-2008$ | 1.1 | 6 | 6 |
| 14 | $2002-2008$ | $2002-2008$ |  | 6 |

TABLE 4.6.2.1.1 Celtic Sea and Division VIIj Herring. CATCH IN NUMBER

```
Units : thousands
    year
age 105195 1959 1960
```



```
    2 3742 25717 72246 16058 18567 51935 15058 70248 19559 39991 54790 93279
    3 33094 2274 24658 32044 19909 13033 17250 9365 59893 20062 39604 55039
    4 25746 19262 3779 5631 48061 4179 6658 15757 9924 49113 11544 33145
    5 12551 11015 13698 2034 8075 20694 1719 3399 13211 9218 22599 12217
    6 55010 34748 19057 14363 21304 9353 12790 25536 21776 26650 15345 28242
    year
age 1970
    1 1319 12658 8422 23547 5507 12768 13317 8159 2800 11335 7162 39361
    2 37260 23313 137690 38133 42808 15429 11113 12516 13385 13913 30093 21285
    3 50087 37563 17855 55805 17184 17783 7286 8610 11948 12399 11726 21861
    4 26481 41904 15842 7012 22530
    5 18763 18759 14531 
    6 19746 21900 11051 12216 8445 7494 9777
        year
age 1982 108198 1984 1985 1986 1987
    1 15339 13540 19517 17916 4159 5976 2307 8260 
    2 42725 102871 92892 57054 56747 67000 82027 42413 41756 63854 26752 94061
    3 8728 26993 41121 36258 42881 43075 30962 68399 24634 38342 35019 9372
    4 4817 3225 16043 16032 32930 23014 9398 19601 35258 16916 27591 10221
    5
    6
    year
age 1994 1995 1996 1997 1998
    1 12130
    2 35768 79159 61923 37440 41510 34072 77378 62153 26472 37006 9470 30710
    361737 22591 38244 53040 27102 36086 18952 35816 18532 24444 46243 5766
    4 3289 36541 7943 31442 28274 14642 12060 5953 5309 14763 21863 18666
    5
    6
    year
age 2006 2007 2008
    1 1460 8043 1288
    2 33894 11028 12468
    3 10914 36223 8144
    4 2469 5509 15565
    5}66261 1365 2328
    6 2997 2509 909
```

TABLE 4.6.2.1.2 Celtic Sea and Division VIIj Herring. WEIGHTS AT AGE IN THE CATCH

```
Units : kg
    year
age 1958
    1 0.096 0.087 0.093 0.098 0.109 0.103 0.105 0.103 0.122 0.119 0.119 0.122
    2 0.115 0.119 0.122 0.127 0.146 0.139 0.139 0.143 0.154 0.158 0.166 0.164
    3 0.162 0.166 0.156 0.156 0.170 0.194 0.182 0.180 0.191 0.185 0.196 0.200
    4 0.185 0.185 0.191 0.185 0.187 0.205 0.215 0.212 0.212 0.217 0.215 0.217
    5 0.205 0.200 0.205 0.207 0.210 0.217 0.225 0.232 0.237 0.243 0.235 0.237
    6 0.224 0.220 0.222 0.224 0.234 0.241 0.235 0.249 0.250 0.257 0.257 0. 252
    year
age 1970
    1 0.128 0.117 0.132 0.125 0.141 0.137 0.137 0.134 0.127 0.127 0.117 0.115
    2 0.162 0.166 0.170 0.174 0.180 0.187 0.174 0.185 0.189 0.174 0.174 0.172
    3 0.200 0.200 0.194 0.205 0.210 0.215 0.205 0.212 0.217 0.212 0.207 0.210
    4 0.225 0.225 0.220 0.215 0.225 0.240 0.235 0.222 0.240 0.230 0.237 0.245
    5 0.240 0.245 0.245 0.245 0.237 0.251 0.259 0.243 0.279 0.253 0.259 0.267
    6 0.262 0.261 0.265 0.269 0.264 0.269 0.278 0.271 0.288 0.282 0.273 0.287
    year
age 198198
    1 0.115 0.109 0.093 0.104 0.112 0.096 0.097 0.106 0.099 0.092 0.096 0.092
    2 0.154 0.148 0.142 0.140 0.155 0.138 0.132 0.129 0.137 0.128 0.123 0.129
    3 0.194 0.198 0.185 0.170 0.172 0.186 0.168 0.151 0.153 0.168 0.150 0.155
    4 0.237 0.220 0.213 0.201 0.187 0.192 0.203 0.169 0.167 0.182 0.177 0.180
    5 0.262 0.276 0.213 0.234 0.215 0.204 0.209 0.194 0.188 0.190 0.191 0.201
    6 0.279 0.305 0.249 0.256 0.252 0.245 0.224 0.208 0.214 0.219 0.205 0.211
    year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 
    1 0.097 0.088 0.088 0.093 0.099 0.090 0.092 0.082 0.096 0.089 0.080 0.077
    2 0.135 0.126 0.118 0.124 0.121 0.120 0.111 0.107 0.115 0.102 0.130 0.102
    3 0.168 0.151 0.147 0.141 0.153 0.149 0.148 0.139 0.139 0.128 0.134 0.142
    4 0.179 0.178 0.159 0.157 0.163 0.167 0.168 0.162 0.156 0.146 0.151 0.147
    5 0.190 0.188 0.185 0.172 0.173 0.180 0.185 0.177 0.185 0.165 0.159 0.158
    6 0.214 0.210 0.210 0.198 0.194 0.191 0.193 0.194 0.201 0.191 0.186 0.174
    year
age 2006 2007 2008
    1 0.093 0.074 0.091
    2 0.105 0.106 0.120
    3 0.127 0.123 0.144
    4 0.151 0.141 0.156
    5 0.155 0.166 0.172
    6 0.168 0.164 0.193
```

TABLE 4.6.2.1.3 Celtic Sea and Division VIIj Herring. WEIGHTS AT AGE IN THE STOCK

```
Units : kg
    year
age 1958 1959 1960 1961 1962 1963 1964 1965 1966 1967 1968 1969
    1 0.096 0.087 0.093 0.098 0.109 0.103 0.105 0.103 0.122 0.119 0.119 0.122
    2 0.115 0.119 0.122 0.127 0.146 0.139 0.139 0.143 0.154 0.158 0.166 0.164
    3 0.162 0.166 0.156 0.156 0.170 0.194 0.182 0.180 0.191 0.185 0.196 0.200
    4 0.185 0.185 0.191 0.185 0.187 0.205 0.215 0.212 0.212 0.217 0.215 0.217
    5 0.205 0.200 0.205 0.207 0.210 0.217 0.225 0.232 0.237 0.243 0.235 0.237
    6 0.224 0.220 0.222 0.224 0.234 0.241 0. 235 0.249 0.250 0.257 0. 257 0. 252
        year
```



```
    1 0.128 0.117 0.132 0.125 0.141 0.137 0.137 0.134 0.127 0.127 0.117 0.115
    2 0.162 0.166 0.170 0.174 0.180 0.187 0.174 0.185 0.189 0.174 0.174 0.172
    3 0.200 0.200 0.194 0.205 0.210 0.215 0.205 0.212 0.217 0.212 0.207 0.210
    4 0.225 0.225 0.220 0.215 0.225 0.240 0.235 0.222 0.240 0.230 0.237 0.245
    5 0.240 0.245 0.245 0.245 0.237 0.251 0.259 0.243 0.279 0.253 0.259 0.267
    6 0.262 0.261 0.265 0.269 0.264 0.269 0.278 0.271 0.288 0.282 0.273 0.287
    year
age 191982 1983 1984 1985
    1 0.115 0.109 0.093 0.104 0.112 0.096 0.097 0.106 0.099 0.092 0.096 0.092
    2 0.154 0.148 0.142 0.140 0.155 0.138 0.132 0.129 0.137 0.128 0.123 0.129
    3 0.194 0.198 0.185 0.170 0.172 0.186 0.168 0.151 0.153 0.168 0.150 0.155
    4 0.237 0.220 0.213 0.201 0.187 0.192 0.203 0.169 0.167 0.182 0.177 0.180
    5 0.262 0.276 0.213 0.234 0.215 0.204 0.209 0.194 0.188 0.190 0.191 0.201
    6 0.279 0.305 0.249 0.256 0.252 0.245 0.224 0.208 0.213 0.219 0.205 0.211
    year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
    1 0.097 0.088 0.088 0.093 0.099 0.090 0.092 0.082 0.096 0.078 0.077 0.074
    2 0.135 0.126 0.118 0.124 0.121 0.120 0.111 0.107 0.115 0.100 0.127 0.103
    3 0.168 0.151 0.147 0.141 0.153 0.149 0.148 0.139 0.139 0.130 0.133 0.145
    4 0.179 0.178 0.159 0.157 0.163 0.167 0.168 0.162 0.156 0.141 0.151 0.143
    5 0.190 0.188 0.185 0.172 0.173 0.180 0.185 0.177 0.184 0.156 0.156 0.155
    6 0.214 0.210 0.210 0.198 0.194 0.191 0.193 0.194 0.201 0.168 0.187 0.167
        year
age 2006 2007 2008
    1 0.085 0.066 0.083
    2 0.104 0.102 0.117
    3 0.123 0.116 0.140
    40.153 0.135 0.156
    5 0.150 0.151 0.170
    6 0.159 0.160 0.180
```


## TABLE 4.6.2.1.4 Celtic Sea and Division VIIj Herring. NATURAL MORTALITY

```
Units : NA
    year
age 1958 1959 1960 1961 1962 1963 1964 1965 1966 1967 1968 1969 1970 1971 1972
    1 1.0
    2 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3
    3
    4
```



```
    % 0.1
age 1973 1974 1975 1976 1977 1978 1979 1980 1981 1982 1983 1984 1985 1986 1987
```



```
    2 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.3 0.0.3 0.0.3 0.3 0.3 0.3 0.3 0.3
    3 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2 0.2
    4}00.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 1 0.1 1 0.1 1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 
    5
    6 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1
    year
age 1988 1989 1990 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002
    1}1.
    1
```



```
    4 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 
    5 0.1 0.1 0.1 0.1 0.1 0.1 0.1 0.1 
    6 0.1 0.1 0.1 0.1 0.1 0.1 0.1 
    year
age 2003 2004 2005 2006 2007 2008
```



```
    2
    3
```



```
    5
    6
```

TABLE 4.6.2.1.5 Celtic Sea and Division VIIj Herring. PROPORTION MATURE

Units : NA
year
age 195819591960196119621963196419651966196719681969197019711972
$\begin{array}{lllllllllllllll}1 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 \\ 0.5\end{array}$
$\begin{array}{lllllllllllllllll}2 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0\end{array}$
$\begin{array}{lllllllllllllll}3 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 1.0\end{array}$
$\begin{array}{llllllllllllllll}4 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0\end{array}$
$\begin{array}{lllllllllllllll}5 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 6 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 1.0\end{array}$ year
age $197319741975197619771978197919801981 \quad 198219831984198519861987$

| 1 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

$\begin{array}{lllllllllllllllll}2 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0\end{array}$
$\begin{array}{lllllllllllllll}3 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 1.0\end{array}$
$\begin{array}{lllllllllllllll}4 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 1.0\end{array}$
$\begin{array}{lllllllllllllll}5 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 6 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 1.0\end{array}$ year
age 198819891990199119921993199419951996199719981999200020012002
$\begin{array}{lllllllllllllll}1 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 & 0.5 \\ 0.5\end{array}$
$\begin{array}{llllllllllllllll}2 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0\end{array}$
$\begin{array}{lllllllllllllll}3 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 1.0\end{array}$
$\begin{array}{lllllllllllllll}4 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 1.0\end{array}$
$\begin{array}{lllllllllllllll}5 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 6 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 1.0\end{array}$ year
age 200320042005200620072008

| 1 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 |
| :--- | :--- | :--- | :--- | :--- | :--- |

$2 \quad 1.0 \quad 1.0 \quad 1.0 \quad 1.0 \quad 1.0 \quad 1.0$
$\begin{array}{lllllll}3 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0\end{array}$
$\begin{array}{lllllll}4 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0\end{array}$
$\begin{array}{lllllll}5 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 \\ 6 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0 & 1.0\end{array}$

TABLE 4.6.2.1.6 Celtic Sea and Division VIIj Herring. FRACTION OF HARVEST BEFORE SPAWNING

```
Units : NA
```

year
age 195819591960196119621963196419651966196719681969197019711972
$\begin{array}{lllllllllllll}1 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 \\ 0.2 & 0.2 & 0.2\end{array}$
$\begin{array}{lllllllllllllllll}2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2\end{array}$
$300.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2$

| 4 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 5 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |

$6 \quad 0.20 .20 .2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2$ year
age $197319741975 \quad 197619771978 \quad 197919801981 \quad 198219831984198519861987$

| 1 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | $0.0 .2 \quad 0.2$

$\begin{array}{llllllllllllllll}2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2\end{array}$

| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | $0.0 .2 \quad 0.2$

$\begin{array}{lllllllllllllll}4 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 \\ 5 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 \\ 0.2\end{array}$
$\begin{array}{llllllllllllllll}6 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2\end{array}$ year
age 198819891990199119921993199419951996199719981999200020012002
$10.20 .2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2$
$20.20 .2 \begin{array}{lllllllllllllll}2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2\end{array}$

| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | $0.2 .2 \quad 0.2$

$4 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2 \quad 0.2$
$\begin{array}{lllllllllllllll}5 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 \\ 6 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2 & 0.2\end{array} 0.2$
year
age $2003 \quad 2004 \quad 2005 \quad 2006 \quad 2007 \quad 2008$
10.5510 .5510 .5510 .5510 .5510 .551
$20.5510 .551 \quad 0.551 \quad 0.5510 .5510 .551$
30.5510 .5510 .5510 .5510 .5510 .551
$40.5510 .5510 .551 \quad 0.5510 .5510 .551$
$50.5510 .551 \quad 0.551 \quad 0.5510 .5510 .551$
$60.5510 .5510 .551 \quad 0.5510 .5510 .551$

TABLE 4.6.2.1.7 Celtic Sea and Division VIIj Herring. FRACTION OF NATURAL MORTALITY BEFORE SPAWNING

```
Units : NA
    year
age 1958 1959 1960 1961 1962 1963 1964 1965 1966 1967 1968 1969 1970 1971 1972
    1 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5
    2 0.5
    3}00.
    4 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5
    5
```



```
    year
age 1973 1974 1975 1976 1977 1978 1979 1980 1981 1982 1983 1984 1985 1986 1987
```



```
    2 0.5 0.5 0.5 0.5 0.5 0.5
    3 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5
    4}00.
    6 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5
    year
age 1988 1989 1990 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002
    1 0.5 0.5 0.5 0.5
    2 0.5
    3 0.5 0.5 0.5 0.5 0.5
    4 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5 0.5
    5 [10.5
    year
age 2003 2004 2005 2006 2007 2008
    1}00.
    2
    3
    4}00.
    5
    6
```

TABLE 4.6.2.1.8 Celtic Sea and Division VIIj Herring. SURVEY INDICES


TABLE 4.6.2.1.9 Celtic Sea and Division VIIj Herring. STOCK OBJECT CONFIGURATION

| min | max plusgroup | minyear | maxyear | minfbar | maxfbar |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 6 | 6 | 1958 | 2008 | 2 |

TABLE 4.6.2.1.10 Celtic Sea and Division VIIj Herring. FLICA CONFIGURATION SETTINGS

| sep. 2 | NA |
| :---: | :---: |
| sep.gradual | TRUE |
| sr | FALSE |
| sr.age | 1 |
| lambda.age | 0.1111110 |
| lambda.yr | 111111 |
| lambda.sr | 0 |
| index.model | linear |
| index.cor | 1 |
| sep.nyr | 6 |
| sep.age | 3 |
| sep.sel | 1 |

## TABLE 4.6.2.1.11 Celtic Sea and Division VIIj Herring. FLR, R SOFTWARE VERSIONS

R version 2.8.0 (2008-10-20)

```
Package : FLICA
Version : 1.4-10
Packaged : Sat Mar 21 18:30:56 2009; mpa
Built : R 2.8.0; ; 2009-03-21 18:30:58; windows
Package : FLAssess
Version : 1.99-102
Packaged : Mon Mar 23 08:18:19 2009; mpa
Built : R 2.8.0; i386-pc-mingw32; 2009-03-23 08:18:21; windows
Package : FLCore
Version : 3.0
Packaged : Tue Mar 10 04:42:26 2009; theussl
Built : R 2.8.1; i386-pc-mingw32; 2009-03-10 04:42:28; windows
```

TABLE 4.6.2.1.12 Celtic Sea and Division VIIj Herring. STOCK SUMMARY
$\left.\begin{array}{lrrrrrr}\text { Year Recruitment } & \text { TSB } & \text { SSB } & \text { Fbar } & \text { Landings } & \text { Landings } \\ & \text { Age } 1 & & & \text { (Ages } 2-5) & \text { f } & \text { tonnes }\end{array}\right]$ SOP

TABLE 4.6.2.1.13 Celtic Sea and Division VIIj Herring. ESTIMATED FISHING MORTALITY

```
Units : f
    year
age 1958
    1 0.009 0.002 0.024 0.015 0.002 0.002 0.011 0.000 0.016 0.017 0.023 0.034
    2 0.169 0.320 0.300 0.314 0.311 0.339 0.182 0.230 0.199 0. 201 0. 0. 279 0.437
    3 0.400 0.155 0.624 0.223 0.877 0.401 0.190 0.174 0.334 0.342 0.333 0.538
    4 0.496 0.407 0.392 0.264 0.573 0.425 0.349 0.252 0.267 0.476 0.321 0.487
    5 0.384 0.362 0.502 0.337 0.649 0.460 0.276 0.268 0.308 0.377 0.371 0.582
    6 0.384 0.362 0.502 0.337 0.649 0.460 0.276 0.268 0.308 0.377 0.371 0.582
    year
age 1970
    1 0.010 0.024 0.051 0.135 0.069 0.148 0.127 0.078 0.033 0.078 0.079 0.164
    2 0.315 0.416 0.710 0.634 0.733 0.504 0.325 0.292 0.309 0.404 0.552 0.657
    3 0.477 0.653 0.709 0.779 0.724 0.869 0.510 0.483 0.537 0.565 0.773 1.153
    4 0.512 0.906 0.607 0.643 0.816 0.756 1.017 0.825 0.633 0.913 0.636 1.020
    5 0.499 0.738 0.832 0.824 0.916 0.814 0.673 0.585 0.553 0.704 0.772 1.078
    6 0.499 0.738 0.832 0.824 0.916 0.814 0.673 0.585 0.553 0.704 0.772 1.078
    year
age 1982 1983}101984 1985 1986 1987 1988 1989 1990 1991 1992 1993
    1 0.037 0.029 0.055 0.056 0.012 0.010 0.009 0.028 0.010 0.017 0.017 0.008
    2 0.489 0.676 0.510 0.398 0.449 0.497 0.308 0.414 0.331 0.626 0.628 0.372
    3 0.678 0.718 0.693 0.410 0.639 0.805 0.486 0.488 0.485 0.621 0.951 0.505
    4 0.826 0.544 1.298 0.609 0.768 0.823 0.381 0.620 0.476 0.691 1.274 0.786
    5 0.763 0.797 0.927 0.555 0.710 0.810 0.457 0.593 0.500 0.781 1.073 0.626
    6 0.763 0.797 0.927 0.555 0.710 0.810 0.457 0.593 0.500 0.781 1.073 0.626
    year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
    1 0.028 0.022 0.016 0.016 0.039 0.047 0.037 0.064 0.012 0.029 0.035 0.030
    2 0.416 0.447 0.341 0.426 0.436 0.606 0.713 0.615 0.248 0.301 0.366 0.316
    3 0.481 0.544 0.434 0.593 0.681 0.936 0.908 0.968 0.399 0.519 0.631 0.544
    4 0.315 0.556 0.353 0.733 0.702 0.957 0.934 0.787 0.336 0.557 0.677 0.584
    5 0.497 0.611 0.450 0.670 0.695 0.957 1.002 0.920 0.380 0.519 0.631 0.544
    6 0.497 0.611 0.450 0.670 0.695 0.957 1.002 0.920 0.380 0.519 0.631 0.544
    year
age 2006 2007 2008
    1 0.017 0.014 0.008
    2 0.176 0.143 0.080
    3 0. 303 0.246 0.137
    4 0.326 0.265 0.147
    5 0.303 0.246 0.137
    6 0.303 0.246 0.137
```

TABLE 4.6.2.1.14 Celtic Sea and Division VIIj Herring. ESTIMATED POPULATION ABUNDANCE

| year |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 |
| 1 | 295674 | 872952 | 190690 | 219996 | 565338 | 284233 | 1081346 | 339817 | 698800 | 714161 |
| 2 | 27804 | 107818 | 320441 | 68501 | 79694 | 207527 | 104391 | 393427 | 124979 | 252952 |
| 3 | 109932 | 17402 | 57996 | 175899 | 37084 | 43241 | 109579 | 64481 | 231565 | 75898 |
| 4 | 68932 | 60308 | 12199 | 25437 | 115175 | 12634 | 23708 | 74183 | 44357 | 135785 |
| 5 | 41209 | 37992 | 36316 | 7456 | 17674 | 58733 | 7472 | 15139 | 52172 | 30721 |
| 6 | 180616 | 119850 | 50523 | 52653 | 46628 | 26545 | 55596 | 113737 | 85997 | 88816 |
|  | year |  |  |  |  |  |  |  |  |  |
| age | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 |
| 1 | 839600 | 444838 | 215365 | 857772 | 264810 | 291386 | 129813 | 144961 | 175297 | 169900 |
| 2 | 258309 | 301787 | 158150 | 78461 | 308204 | 92537 | 93642 | 44569 | 45987 | 56818 |
| 3 | 153271 | 144704 | 144445 | 85460 | 38340 | 112301 | 36363 | 33321 | 19951 | 24615 |
| 4 | 44120 | 89909 | 69195 | 73373 | 36400 | 15449 | 42175 | 14432 | 11444 | 9808 |
| 5 | 76349 | 28974 | 49965 | 37538 | 26842 | 17949 | 7347 | 16883 | 6130 | 3743 |
| 6 | 51842 | 66979 | 52583 | 43823 | 20413 | 22719 | 14684 | 14048 | 20867 | 8961 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 1 | 134846 | 238148 | 148152 | 405081 | 671728 | 743311 | 572200 | 516009 | 537975 | 976622 |
| 2 | 57785 | 47982 | 81055 | 50361 | 126419 | 238215 | 265588 | 199194 | 179450 | 195492 |
| 3 | 31437 | 31419 | 23735 | 34592 | 19350 | 57458 | 89722 | 118099 | 99124 | 84819 |
| 4 | 12437 | 15041 | 14627 | 8973 | 8940 | 8046 | 22946 | 36734 | 64160 | 42826 |
| 5 | 3888 | 5973 | 5459 | 7007 | 2929 | 3540 | 4227 | 5667 | 18071 | 26941 |
| 6 | 8259 | 7826 | 10125 | 8541 | 8788 | 3686 | 3230 | 1519 | 2603 | 8748 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| 1 | 393365 | 475553 | 429569 | 180747 | 959501 | 330034 | 702463 | 683183 | 341631 | 372002 |
| 2 | 355804 | 143369 | 170150 | 156458 | 65382 | 346931 | 120478 | 251378 | 245839 | 123659 |
| 3 | 88076 | 193785 | 70214 | 90538 | 61992 | 25849 | 177099 | 58896 | 119096 | 129472 |
| 4 | 31044 | 44365 | 97368 | 35411 | 39844 | 19609 | 12768 | 89669 | 27995 | 63208 |
| 5 | 17016 | 19182 | 21601 | 54710 | 16048 | 10087 | 8087 | 8434 | 46551 | 17801 |
| 6 | 12268 | 18411 | 17662 | 17342 | 44408 | 22651 | 23165 | 20072 | 17897 | 18661 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 1 | 241493 | 491902 | 434874 | 401671 | 479767 | 105907 | 257960 | 760084 | 265482 | 585465 |
| 2 | 134615 | 85465 | 172685 | 154206 | 138663 | 174464 | 37863 | 91655 | 271355 | 96045 |
| 3 | 59842 | 64513 | 34532 | 62732 | 61747 | 80164 | 95656 | 19453 | 49516 | 168577 |
| 4 | 58557 | 24785 | 20716 | 11405 | 19514 | 33924 | 39075 | 41686 | 9244 | 29933 |
| 5 | 27477 | 26256 | 8613 | 7365 | 4696 | 12623 | 17593 | 17969 | 21035 | 6040 |
| 6 | 15440 | 22516 | 16118 | 6537 | 6835 | 17139 | 4808 | 6220 | 12007 | 12045 |
| year |  |  |  |  |  |  |  |  |  |  |
| age 2008 |  |  |  |  |  |  |  |  |  |  |
| 1360168 |  |  |  |  |  |  |  |  |  |  |
| 2212473 |  |  |  |  |  |  |  |  |  |  |
| 361669 |  |  |  |  |  |  |  |  |  |  |
| 4107869 |  |  |  |  |  |  |  |  |  |  |
| 5 | 20788 |  |  |  |  |  |  |  |  |  |
| 6 | 7442 |  |  |  |  |  |  |  |  |  |

TABLE 4.6.2.1.15 Celtic Sea and Division VIIj Herring. SURVIVORS AFTER TERMINAL YEAR

```
Units : NA
    year
age 2009
            NA
    2 98648
    3145357
    444017
    5 84239
    6 22269
```


## TABLE 4.6.2.1.16 Celtic Sea and Division VIIj Herring. FITTED SELECTION PATTERN

```
Units : NA
    year
age}20003 2004 2005 2006 2007 2008
    1 0.055 0.055 0.055 0.055 0.055 0.055
    2 0.580 0.580 0.580 0.580 0.580 0.580
    31.000 1.000 1.000 1.000 1.000 1.000
    4 1.073 1.073 1.073 1.073 1.073 1.073
    5 1.000 1.000 1.000 1.000 1.000 1.000
    6 1.000 1.000 1.000 1.000 1.000 1.000
```

TABLE 4.6.2.1.17 Celtic Sea and Division VIIj Herring. PREDICTED CATCH IN NUMBERS

```
Units : NA
    year
age 1958 1959 1960 1961 1962 1963 1964 1965 1966 1967 1968 1969
```



```
    2 3742 25717 72246 16058 18567 51935 15058 70248 19559 39991 54790 93279
    3 33094 2274 24658 32044 19909 13033 17250 9365 59893 20062 39604 55039
    4 25746 19262 3779 5631 48061 4179 6658 15757 9924 49113 11544 33145
    5 12551 11015 13698 2034 8075 20694 1719 3399 13211 9218 22599 12217
    6 55010 34748 19057 14363 21304 9353 12790 25536 21776 26650 15345 28242
        year
age 1970 1971 1972 1973 1974 1975 1976 1977 1978 1979 1980
    1 1319 12658 8422 23547 5507 12768 13317 8159 2800 11335 7162 39361
    2 37260 23313 137690 38133 42808 15429 11113 12516 13385 13913 30093 21285
    3 50087 37563 17855 55805 17184 17783 7286 8610 11948 12399 11726 21861
    4 26481 41904 15842 7012 22530
    5 18763 18759 14531 9651 4225 
```



```
    year
age 1982 1983 1984 1985 1986 1987 1988
    1 15339 13540 19517 17916 4159 5976 2307 8260 
    2 42725 102871 92892 57054 56747 67000 82027 42413 41756 63854 26752 94061
    3 8728 26993 41121 36258 42881 43075 30962 68399 24634 38342 35019 9372
    4 4817 
    5
    6
        year
age 1994 1995 1996 1997 1998 1999 2000 2001 2002 
    1 12130
    2 35768 79159 61923 37440 41510 34072 77378 62153 26472 39467 10117 21609
    3 61737 22591 38244 53040 27102 36086 18952 35816 18532 29653 40974 7465
    4 3289 36541 7943 31442 28274 14642 12060 5953 5309 13844 18390 17632
    5 3025
    6
        year
age 2006 2007 2008
    1 }2027875001 128
    2 38007 11099 14077
    3 11801 33514 7182
    4 2451 6636 14074
    5}55251 1258 2539
    6 2997 2509 909
```

TABLE 4.6.2.1.18 Celtic Sea and Division VIIj Herring. CATCH RESIDUALS

| Units <br> year |  |  |  |  |  |  |  | thousands NA |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |  |  |  |  |  |  |  |
| 1 | 0.36 | NA | NA | NA | NA | NA |  |  |  |  |  |  |  |
| 2 | NA | NA | NA | NA | NA | NA |  |  |  |  |  |  |  |
| 3 | NA | NA | NA | NA | NA | NA |  |  |  |  |  |  |  |
| 4 | NA | NA | NA | NA | NA | NA |  |  |  |  |  |  |  |
| 5 | NA | NA | NA | NA | NA | NA |  |  |  |  |  |  |  |
| 6 | NA | NA | NA | NA | NA | NA |  |  |  |  |  |  |  |

TABLE 4.6.2.1.19 Celtic Sea and Division VIIj Herring. PREDICTED INDEX VALUES
Celtic Sea Herring Acoustic

| Units : |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| year |  |  |  |  |  |  |  |
| y NA NA |  |  |  |  |  |  |  |
| age | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| 2 | 121 | 144 | NA | 75 | 255 | 93 | 220 |
| 3 | 80 | 92 | NA | 22 | 71 | 254 | 104 |
| 4 | 23 | 32 | NA | 39 | 11 | 38 | 154 |
| 5 | 4 | 10 | NA | 14 | 21 | 6 | 25 |

TABLE 4.6.2.1.20 Celtic Sea and Division VIIj Herring. INDEX RESIDUALS
Celtic Sea Herring Acoustic

| Units $: ~ N A ~$ |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| year |  |  |  |  |  |  |  |
| age | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| 2 | 0.425 | -0.851 | NA | 0.606 | -0.190 | -0.286 | 0.295 |
| 3 | 0.633 | -0.422 | NA | 0.258 | -0.390 | -0.145 | 0.067 |
| 4 | 0.249 | -0.629 | NA | 0.340 | 0.198 | -0.207 | 0.048 |
| 5 | 0.417 | -0.633 | NA | 0.425 | -0.648 | 0.343 | 0.096 |

TABLE 4.6.2.1.21 Celtic Sea and Division VIIj Herring. FIT PARAMETERS

|  | Value | Std.dev | Lower.95.pct.CL | Upper.95.pct.CL |
| :---: | :---: | :---: | :---: | :---: |
| F, 2003 | 0.52 | 0.16 | 0.38 | 0.71 |
| F, 2004 | 0.63 | 0.16 | 0.46 | 0.86 |
| F, 2005 | 0.54 | 0.19 | 0.38 | 0.78 |
| F, 2006 | 0.30 | 0.22 | 0.20 | 0.46 |
| F, 2007 | 0.25 | 0.24 | 0.15 | 0.40 |
| F, 2008 | 0.14 | 0.28 | 0.08 | 0.24 |
| Selectivity at age 1 | 0.06 | 0.35 | 0.03 | 0.11 |
| Selectivity at age 2 | 0.58 | 0.14 | 0.44 | 0.77 |
| Selectivity at age 4 | 1.07 | 0.11 | 0.86 | 1.34 |
| Terminal year pop, age 1 | 270188.57 | 0.85 | 50865.16 | 1435203.53 |
| Terminal year pop, age 2 | 212472.30 | 0.32 | 113913.52 | 396304.83 |
| Terminal year pop, age 3 | 61667.80 | 0.26 | 37075.53 | 102572.19 |
| Terminal year pop, age 4 | 107868.04 | 0.25 | 66123.43 | 175966.57 |
| Terminal year pop, age 5 | 20787.44 | 0.26 | 12587.42 | 34329.35 |
| Last true age pop, 2003 | 12621.90 | 0.24 | 7961.68 | 20009.89 |
| Last true age pop, 2004 | 17591.60 | 0.20 | 11851.66 | 26111.48 |
| Last true age pop, 2005 | 17967.65 | 0.21 | 11906.14 | 27115.13 |
| Last true age pop, 2006 | 21033.83 | 0.23 | 13459.60 | 32870.36 |
| Last true age pop, 2007 | 6038.73 | 0.24 | 3803.14 | 9588.47 |
| Index 1, age 2 numbers, Q | 0.00 | 0.24 | 0.00 | 0.00 |
| Index 1, age 3 numbers, Q | 0.00 | 0.24 | 0.00 | 0.00 |
| Index 1, age 4 numbers, Q | 0.00 | 0.25 | 0.00 | 0.00 |
| Index 1, age 5 numbers, Q | 0.00 | 0.27 | 0.00 | 0.00 |

Table 4.7.1.1. Celtic Sea \& Division VIIj Herring. Inputs to the Short Term Forecast

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 360168 | 1 | 0.5 | 0.551 | 0.5 | 0.078 | 7.57E-03 | 0.086 |
| 2 | 131499.8 | 0.3 | 1 | 0.551 | 0.5 | 0.107667 | 7.96E-02 | 0.110 |
| 3 | 145357.2 | 0.2 | 1 | 0.551 | 0.5 | 0.126333 | 0.137198 | 0.131 |
| 4 | 44017.17 | 0.1 | 1 | 0.551 | 0.5 | 0.148 | 0.147265 | 0.149 |
| 5 | 84238.56 | 0.1 | 1 | 0.551 | 0.5 | 0.157 | 0.137198 | 0.164 |
| 6 | 22269.35 | 0.1 | 1 | 0.551 | 0.5 | 0.166333 | 0.137198 | 0.175 |
| 2010 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 1 | 360168 | 1 | 0.5 | 0.551 | 0.5 | 0.078 | 7.57E-03 | 0.086 |
| 2 | . | 0.3 | 1 | 0.551 | 0.5 | 0.107667 | 7.96E-02 | 0.110 |
| 3 | . | 0.2 | 1 | 0.551 | 0.5 | 0.126333 | 0.137198 | 0.131 |
| 4 | - | 0.1 | 1 | 0.551 | 0.5 | 0.148 | 0.147265 | 0.149 |
| 5 | . | 0.1 | 1 | 0.551 | 0.5 | 0.157 | 0.137198 | 0.164 |
| 6 | . | 0.1 | 1 | 0.551 | 0.5 | 0.166333 | 0.137198 | 0.175 |
| 2011 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 1 | 360168 | 1 | 0.5 | 0.551 | 0.5 | 0.078 | 7.57E-03 | 0.086 |
| 2 | . | 0.3 | 1 | 0.551 | 0.5 | 0.107667 | $7.96 \mathrm{E}-02$ | 0.110 |
| 3 | . | 0.2 | 1 | 0.551 | 0.5 | 0.126333 | 0.137198 | 0.131 |
| 4 | . | 0.1 | 1 | 0.551 | 0.5 | 0.148 | 0.147265 | 0.149 |
| 5 | . | 0.1 | 1 | 0.551 | 0.5 | 0.157 | 0.137198 | 0.164 |
| 6 | . | 0.1 | 1 | 0.551 | 0.5 | 0.166333 | 0.137198 | 0.175 |

Table 4.7.1.2. Celtic Sea \& Division VIIj Herring. Single catch option table from the Short Term Forecast

| Year: | 2009 | F multiplier: | 1.0761 | Fbar: | 0.1349 |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | :--- | ---: | ---: | ---: | ---: | ---: |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0081 | 1847 | 159 | 360168 | 28093 | 180084 | 14047 | 108738 | 8482 |
| 2 | 0.0857 | 9349 | 1031 | 131500 | 14158 | 131500 | 14158 | 107964 | 11624 |
| 3 | 0.1476 | 18127 | 2381 | 145357 | 18363 | 145357 | 18363 | 121249 | 15318 |
| 4 | 0.1585 | 6147 | 918 | 44017 | 6515 | 44017 | 6515 | 38369 | 5679 |
| 5 | 0.1476 | 11017 | 1810 | 84239 | 13225 | 84239 | 13225 | 73870 | 11598 |
| 6 | 0.1476 | 2912 | 510 | 22269 | 3704 | 22269 | 3704 | 19528 | 3248 |
| Total |  | 49399 | 6809 | 787550 | 84059 | 607466 | 70012 | 469718 | 55948 |
|  |  |  |  |  |  |  |  |  |  |
| Year: | 2010 | F multiplier: | 1 | Fbar: | 0.1253 |  |  |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0076 | 1717 | 148 | 360168 | 28093 | 180084 | 14047 | 108772 | 8484 |
| 2 | 0.0796 | 8707 | 961 | 131424 | 14150 | 131424 | 14150 | 108262 | 11656 |
| 3 | 0.1372 | 10414 | 1368 | 89418 | 11296 | 89418 | 11296 | 75018 | 9477 |
| 4 | 0.1473 | 13396 | 2000 | 102674 | 15196 | 102674 | 15196 | 90054 | 13328 |
| 5 | 0.1372 | 4152 | 682 | 33991 | 5337 | 33991 | 5337 | 29979 | 4707 |
| 6 | 0.1372 | 10155 | 1777 | 83145 | 13830 | 83145 | 13830 | 73331 | 12197 |
| Total |  | 48541 | 6936 | 800820 | 87902 | 620736 | 73855 | 485417 | 59850 |


| Year: | 2011 | F multiplier: | 1 | Fbar: | 0.1253 |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0076 | 1717 | 148 | 360168 | 28093 | 180084 | 14047 | 108772 | 8484 |
| 2 | 0.0796 | 8712 | 961 | 131500 | 14158 | 131500 | 14158 | 108325 | 11663 |
| 3 | 0.1372 | 10471 | 1375 | 89910 | 11359 | 89910 | 11359 | 75430 | 9529 |
| 4 | 0.1473 | 8327 | 1244 | 63824 | 9446 | 63824 | 9446 | 55979 | 8285 |
| 5 | 0.1372 | 9793 | 1609 | 80181 | 12588 | 80181 | 12588 | 70718 | 11103 |
| 6 | 0.1372 | 11286 | 1975 | 92401 | 15369 | 92401 | 15369 | 81495 | 13555 |
| Total |  | 50307 | 7312 | 817984 | 91014 | 637900 | 76967 | 500719 | 62620 |

Table 4.7.1.3. Celtic Sea \& Division VIIj Herring. Multiple catch option table from the Short Term Forecast

| 2009 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Biomass | SSB | FMult | FBar | Landings |
| 84059 | 55948 | 1.0761 | 0.1349 | 6809 |


| 2010 | 2011 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 87902 | 63606 | 0 | 0 | 0 | 97709 | 72878 |
| - | 63218 | 0.1 | 0.01 | 734 | 97000 | 71759 |
| - | 62833 | 0.2 | 0.03 | 1458 | 96299 | 70662 |
| - | 62451 | 0.3 | 0.04 | 2174 | 95608 | 69586 |
| - | 62071 | 0.4 | 0.05 | 2880 | 94926 | 68531 |
| - | 61694 | 0.5 | 0.06 | 3577 | 94253 | 67496 |
| - | 61320 | 0.6 | 0.08 | 4266 | 93588 | 66482 |
| - | 60949 | 0.7 | 0.09 | 4946 | 92932 | 65488 |
| - | 60580 | 0.8 | 0.10 | 5618 | 92284 | 64513 |
| - | 60214 | 0.9 | 0.11 | 6281 | 91645 | 63557 |
| - | 59850 | 1 | 0.13 | 6936 | 91014 | 62620 |
| - | 59489 | 1.1 | 0.14 | 7583 | 90391 | 61700 |
| - | 59131 | 1.2 | 0.15 | 8221 | 89776 | 60799 |
| - | 58775 | 1.3 | 0.16 | 8852 | 89169 | 59915 |
| - | 58563 | 1.36 | 0.17 | 9227 | 88808 | 59393 |
| - | 58422 | 1.4 | 0.18 | 9475 | 88570 | 59048 |
| - | 58071 | 1.5 | 0.19 | 10090 | 87978 | 58198 |
| - | 57723 | 1.6 | 0.20 | 10697 | 87394 | 57365 |
| - | 57377 | 1.7 | 0.21 | 11297 | 86818 | 56547 |
| - | 57034 | 1.8 | 0.23 | 11889 | 86249 | 55746 |
| - | 56693 | 1.9 | 0.24 | 12474 | 85687 | 54959 |
| - | 56355 | 2 | 0.25 | 13052 | 85132 | 54188 |



Figure 4.1.2.1. Celtic Sea and VIIj herring. Irish official herring catches by statistical rectangle in 2008/2009.


Figure 4.1.3.1 Celtic Sea and Division VIIj - working group estimates of herring landings per season.


Figure 4.2.1.1. Celtic Sea and Division VIIj. Catch numbers at age standardised by yearly mean. 9ringer is the plus group.


Figure 4.2.1.2. Celtic Sea and Division VIIj. Catch numbers at age standardised by yearly mean. 6ringer is the plus group.


Figure 4.2.1.3. Celtic Sea and VIIj herring. The percentage age composition in the survey and the commercial fishery 2008/2009.


Figure 4.3.1.1a Celtic Sea and VIIj herring. Acoustic survey track and haul positions from acoustic survey, October 2008.


Figure 4.3.1.1b. Celtic Sea and VIIj herring. Acoustic survey 2008, total Sa values attributed to herring.


Figure 4.4.1.1. Celtic Sea and VIIj herring. Trends over time in mean weight at age in the catch from 1-9+


Figure..4.4.1.2. Celtic Sea and VIIj herring. Trends over time in mean weight at age in the stock at spawning time from 1-9+


Figure 4.6.1.1. Celtic Sea and VIIj herring. Weighted residuals for FICA assessments using 6+ (left) and 7+ (right).


Figure 4.6.1.2. Celtic Sea and VIIj herring. Catch Diagnostics from FICA exploratory runs with 6+ (left) and 7+(right).


Figure 4.6.1.3. Celtic Sea and VIIj herring. Weighted residuals from FICA exploratory runs with the full survey time series (left) and the 2002-2008 series (right).


Figure 4.6.1.4. Celtic Sea and VIIj herring. Estimates of SSB (above) and Mean F (below) from parametric bootstrapping, for exploratory assessment runs using different settings.


Figure 4.6.1.5. Celtic Sea and VIIj herring. Results of parametric bootstrapping from FLICA for the 6+ assessment (left) and 7+ (right).




Figure 4.6.1.6. Celtic Sea and VIIj herring. Historical Retrospective based on the final assessment in 2009 and spaly runs in 2004-2008.

## Celtic Sea Herring Acoustic, age 2, diagnostics



Figure 4.6.2.1. Celtic Sea and VIIj herring. Diagnostics from the Acoustic survey age 2.

## Celtic Sea Herring Acoustic, age 3, diagnostics



Figure 4.6.2.2. Celtic Sea and VIIj herring

## Celtic Sea Herring Acoustic, age 4, diagnostics



Figure 4.6.2.3. Celtic Sea and VIIj herring. Diagnostics from the Acoustic survey age 4

## Celtic Sea Herring Acoustic, age 5, diagnostics



Figure 4.6.2.4. Celtic Sea and VIIj herring. Diagnostics from the Acoustic survey age 5

## Celtic Sea Herring Retrospective Summary Plot



Figure 4.6.2.5. Celtic Sea and VIIj herring. Analytical Retrospective based on the final assessment

Celtic Sea Herring Retrospective selectivity pattern


Figure 4.6.2.6. Celtic Sea and VIIj herring. Retrospective selection pattern.



| Reference point | F multiplier | Absolute F |
| :--- | :--- | :--- |
| Fbar(2-5) | 1 | 0.1253 |
| FMax | $>=1000000$ |  |
| F0.1 | 1.3466 | 0.1688 |
| F35\%SPR | 1.5231 | 0.1909 |

Figure 4.7.2.1 Celtic Sea and Division VIIj Herring. Yield per Recruit Curve and Short Term Forecast

## 5 West of Scotland Herring

The location of the area occupied by the stock is shown in Figure 5.1. This is an update assessment.

### 5.1 The Fishery

### 5.1.1 ACFM Advice Applicable to 2008 and 2009

ACFM reported in 2008 that the stock was fluctuating at a low level and was being exploited above Fmsy. Recruitment has been low since 1998, and the 2001 and 2002 year classes were very weak.

There was an agreed assessment in 2008. The basis for the advice continued to be based on the proposed management plan. A slightly different plan was accepted by the European Commission in December 2008 (see Section 5.1.3 below). The assessment is considered to be noisy but unbiased. Medium-term evaluations of the proposed management plan had been carried out assuming the same level of noise as seen in the assessment, so management under this plan was considered by ICES in 2005 (and subsequent years) to be precautionary. Fishing according to the proposed management plan would have implied catches up to 13000 tonnes in 2009.

The agreed TAC for 2009 is $21760 t$, which is in accordance with the agreed plan (see Section 5.1.3) but not in accordance with the proposed plan. The TAC in 2008 was 27 200 t .

### 5.1.2 Changes in the Vla (North) Fishery.

Historically, catches have been taken from this area by three fisheries.
i) A Scottish domestic pair trawl fleet and the Northern Irish fleet operated in shallower, coastal areas, principally fishing in the Minches and around the Island of Barra (Figure 5.1) in the south; younger herring are found in these areas. This fleet has reduced in recent years.
ii) The Scottish single boat trawl and purse seine fleets, with refrigerated seawater tanks, targeting herring mostly in the northern North Sea, but also operated in the northern part of VIa (N). This fleet now operates mostly with trawls but many vessels can deploy either gear.
iii) An international freezer-trawler fishery has historically operated in deeper water near the shelf edge where older fish are distributed. These vessels are mostly registered in the Netherlands, Germany, France and England but most are Dutch owned.
In recent years the catch of these last two fleets has become more similar
In 2008, the Scottish trawl fleet fished predominantly in areas similar to the freezer trawler fishery, and hardly in the coastal areas in the southern part of VIa (N). The Northern Irish fleet fished near the Island of Barra. In common with 2006 and 2007, but in contrast to most of the previous years' fisheries, in $200883 \%$ of the fishery was prosecuted in quarter 3 and $99 \%$ of those catches were distributed in the northern part of the area. Prior to 2006 there was a much more even distribution of effort, both temporally and spatially.

### 5.1.3 Regulations and their affects

COUNCIL REGULATION (EC) No 1300/2008 of 18 December 2008 established a multi-annual management agreement for the stock of herring distributed to the west of Scotland and the fisheries exploiting that stock.
$\mathrm{F}=0.25$ if SSB $>75000 \mathrm{t}$
$\mathrm{F}=0.20$ if SSB $<75000 \mathrm{t}$ but $>62500 \mathrm{t}$
$20 \%$ constraint on TAC change.
$\mathrm{F}=0.20$ if $\mathrm{SSB}<62500 \mathrm{t}$ but $>50000 \mathrm{t}$
$25 \%$ constraint on TAC change
$\mathrm{F}=0 \quad$ if $\mathrm{SSB}<50000 \mathrm{t}$.
There is derogation from the above constraints. If STECF considers that the herring stock in the area west of Scotland is failing properly to recover, the TAC constraints may differ from those in the management agreement. This plan is similar but not identical to the proposed plan. The differences and potential impact are discussed below in Section 5.8.

As a result of perceived problems of area misreporting of catch from IVa into VIa (N), Scotland introduced a fishery regulation in 1997 with the aim to improve reporting accuracy. Under this regulation, Scottish vessels fishing for herring were required to hold a license either to fish in the North Sea or in the west of Scotland area (VIa (N)). Only one licensed option could be held at any one time. However in 2004, the requirement to carry only a single licence was rescinded. Area misreporting of catch taken in area IVa into area VIa (N) then increased in 2004 and continued in 2005. It is possible, therefore, that the relaxation of this single area licence contributed to a resurgence in area misreporting at that time. In 2007, as in 2006, there was no misreporting from IVa into VIa (N). New sources of information on catch misreporting from the UK became available in 2006 (see the 2007 HAWG report). This information was associated with a stricter enforcement regime that may have been responsible for the lack of that area misreporting since 2006. However, in 2008 there was again misreporting of some catch from IVa into VIa (N).

The Butt of Lewis box, (a seasonal closure to pelagic fishing of the spawning ground in the north west of the continental shelf in area VIa (N) since the late 1970s (Figure 5.1)) has been opened to fishing following a STECF review in 2007. It has not been possible to show either beneficial or deleterious effects from this closure.

### 5.1.4 Catches in 2008 and Allocation of Catches to Area for VIa (N)

For 2008 the preliminary report of official catches corresponding to the VIa (N) herring stock unit total 25216 t , compared with the TAC of 27200 t . The Working Group's estimates of area misreported and unallocated catches are 9162 t . Discarding is not perceived to be a problem.

The Working Group's best estimate of removals from the stock in 2008 is 16054 t (Table 5.1.1).

### 5.2 Biological composition of the catch

Catch and sample data, by country and by period (quarter), are detailed in Table 5.2.1. The number of samples used to allocate an age-distribution for the VIa (N) catches have continued at the low level seen over the last few years (except in 2006). There were 13 samples available in 2008, obtained from the Scottish (11), Dutch (1) and English (1) fleets. The Dutch and English fleets each took a similar magnitude of catches in the area, slightly less than half the Scottish catches. The samples available
were used to allocate a mean age-structure (using the sample number weighting) to unsampled catches, in the same or adjacent quarters, as no sampling data were available for other quarters. The allocation of age structures to unsampled catches, and the calculation of total international catch-at-age and mean weight-at-age in the catches were made using the 'sallocl' programme (Patterson, 1998a). As 11 of the 13 samples obtained came from only one of the major fisheries in one quarter (Scotland $3^{\text {rd }}$ quater); it is likely that they are reasonably representative of these catches, but do not fully reflect the entire fishery.

Catch number- and weight-at-age information is given in the ICA stock report section 5.6 (cf Table 5.6.1.1 and 5.6.1.2 respectively). Two larger year classes can be seen clearly in the catch-at-age table: 2000 and 2004 at 7-and 3-ringers respectively in 2008. The 2001, 2002 and 2003 year classes all appear relatively weak, with the 2002 year class the weakest. 1-ring herring in the catch are observed intermittently and are rarely representative of year class strength and are down-weighted in the assessment, (see Section 5.6).

### 5.3 Fishery Independent Information

### 5.3.1 Acoustic Survey

The 2008 acoustic survey was carried out from the $27^{\text {th }}$ June to the $16^{\text {th }}$ July 2008 using a chartered commercial fishing vessel (MFV Chris Andra). Further details are available in the Report of the Planning Group for International Pelagic Surveys (ICES 2009/LRC:02). The commercial vessel changes through the time series, though year effects seen in the series are not linked to vessel effects. The biomass estimate for VIa $(\mathrm{N})$ from the acoustic survey (Table 5.3.1) has increased by approximately $165 \%$ from 2007 (from 298880 tonnes to 791350 tonnes), to give the second highest estimate in the time series. The estimate has increased due to increased numbers, but also mean weights (see section 5.4). In 2007 very few fish below 20 cm or above 31 cm were seen giving a weight/length relationship that had a lower gradient than in previous years. The survey catches in 2008 gave a wider spread ( 15 to 35 cm ) resulting in a more representative weight/length relationship.

In 2008 quite similar year class proportions were seen in the catch and the survey. However, the survey showed slightly higher proportions of 4- and 5-ringers, whereas the catch showed higher proportions of 2-ring fish. There is no basis for concluding which of the sources of data are more reliable, the catch is sparsely and partially sampled and the survey in 2008 appeared to catch fewer 2-ring herring (ICES 2009/LRC:02) (cf. Figure 5.6.2.12 for residuals in the fitted model).

### 5.4 Mean weights-at-age and maturity-at-age

### 5.4.1 Mean Weight-at-age

Weights-at-age in the stock from acoustic surveys are given in Table 5.3.1 and weights-at-age in the catches are given in Section 5.6.2 (cf. Table 5.6.2.2) and are used in the assessment. The weights-at-age in the catch are comparable to previous years for older ages, with slightly higher weights from 3- to 6-ring herring. The weights-atage in the stock are, again, normal for the older ages but slightly higher than normal for the 2 -ring herring (cf. Table 5.6.1.3). This is likely a reflection of the more representative catch in the survey in 2008 of older ages and lower than normal catch of 2ring herring (ICES 2009/LRC:02).

### 5.4.2 Maturity Ogive

The maturity ogive is obtained from the acoustic survey (Table 5.3.1). The survey provides estimated values for the period 1987 and 1992 to 2008 (cf. Table 5.6.2.5). In $2008,98 \%$ of the 2 -ring fish caught were mature, this is the second highest proportion mature at this age since 1992 when measurements began, with the highest value (virtually $100 \%$ mature) seen in 2007. The sensitivity of the assessed SSB to the estimated maturity was investigated in 2008 where the assessment was re-run with fraction mature at 2-ring taken from average maturity for the years 2004-2006. This resulted in a $4 \%$ reduction of SSB in 2007. This was considered to be negligible in the context of the precision of the estimate of SSB.

### 5.5 Recruitment

There are no specific recruitment indices for this stock. Although both catch and acoustic survey generally have some catches at 1-ring both the fishery and survey encounter this age group only incidentally. The first reliable appearance of a cohort appears at 2 -ring in both the catch and the stock. Thus in predictions, estimates of both 1- and 2-ring herring numbers from the assessment are replaced for prediction years.

### 5.6 Assessment of VIa (North) herring

### 5.6.1 Data Exploration and Preliminary Modelling

The ICA assessment (Patterson 1998a), implemented in FLR (Kell 2007) as FLICA, has exhibited substantial revision both up and down over the last few years, largely due to the noisy survey used for tuning the assessment. The model settings were last explored in detail in 2003 (ICES 2003/ACFM:17). In order to establish if different model settings would give improved consistency in the assessment and subsequent advice, the settings of the model were explored particularly with respect to retrospective performance.

Range of model settings evaluated:

- Selection at oldest true age from 0.7 to 1.2 (current value is 1.0 )
- Balance of Survey and Catch weighting by varying catch weighting from 0.5 to 2.0 (current value is 1.0 )
- Weighting at age: flat or inverse variance (with 1-ring herring down weighted). (Current weightings for catch and survey are flat with 1-ring herring downweighted).
Weightings at age

| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | $9+$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catch flat | 0.1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| Catch inv var at age | 0.10 | 3.67 | 2.87 | 2.23 | 1.74 | 1.37 | 1.04 | 0.94 | 0.91 |
| Survey flat | 0.1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| Survey inv var at age | 0.10 | 0.83 | 1.26 | 1.21 | 1.14 | 1.22 | 1.13 | 1.06 | 0.87 |

The age dependent weights for catch sampling were taken from the North Sea as no analysis was available. The VIa ( N ) acoustic survey was analysed for variance at age and the mean over all years from 1992 computed. Earlier surveys had little data on numbers at age and did not give information on variance at age.

Two measures of retrospective performance were selected.

1. Mean deviance between previous assessments and most recent assessment.

$$
\sum_{r=1}^{n}\left(S_{r, t-r-y}-S_{f, t-r-y}\right) / n
$$

where $y$ runs from 1 to $n$.
2. Mean square deviance between previous assessments and most recent assessment.

$$
\sum_{r=1}^{n}\left(S_{r, t-r-y}-S_{f, t-r-y}\right)^{2} / n
$$

Where $S_{r, t-r-y}$ is SSB in the retrospective run $r$ and year $t-r-y$, where $t$ is the terminal year in the most recent assessment and $r$ the number of years gone back in the retrospective and $y$ the number of years back from the last year in the assessment; $S_{f, t-r-y}$ is the equivalent term for the $\mathrm{SSB}_{f}$ as estimated in the most recent assessment $f$. For F or R substitute for S .

Figures 5.6.1.1 to 5.6.1.4 show the range of squared deviance plotted against y for the two different types of weightings using two terminal years for data, 2008 and 2009 WG data.

The optimal choice of setting is selected from the assessment that exhibits the minimum overall squared deviation. These are illustrated for different terminal years $(2008,2009)$ and the two different data set weighting methods (Flat or Age based) in Figures 5.6.1.1 - 5.6.1.4 and optimal choices are:

| Model settings | Optimal choices |  |  |  |
| :--- | :---: | :---: | :---: | :---: |
| Catch and survey weighting <br> method | Data <br> sets | Selection at oldest <br> age | Weight on <br> Catch | Weight on <br> survey |
| Flat | 2009 | 0.8 | $85 \%$ | $15 \%$ |
| Inverse var at age | 2009 | 0.9 | $82 \%$ | $18 \%$ |
| Flat | 2008 | 1.0 | $69 \%$ | $31 \%$ |
| Inverse var at age | 2008 | 1.1 | $77 \%$ | $23 \%$ |
| SPALY | 2009 | 1.0 | $81 \%$ | $19 \%$ |

Based on the value of the mean squared retrospective deviation the best option for both 2008 and 2009 is the age weighted method, but the level of weight given to survey / catch is heavily dependent on the terminal year used to do the analysis. The selection at oldest age is not consistent across years or choice of weighting. As the results depend strongly on the year the data is examined there is no basis for changing the selection. A comparison between retrospective performance based on SPALY settings and the optimal 2009 settings (Figure 5.6 .1 .5 ) shows only modest improvement, which would be cancelled out had we chosen the 2008 optimal settings for 2009 data. Although it would appear to be better to change to age weighting it is very dif-
ficult to establish the values to use and there is no reason to believe that choosing the optimal method from 2009 by this method would hold for 2010. Fit to the catch is more or less identical (Figure 5.6.1.5b). On this basis there seems little justification in proposing a different approach.

We conclude that continuing with the current weighting and model settings is an acceptable solution, until more data, possibly as a result of the extended surveys from SGHERWAY, are available.

### 5.6.2 Stock Assessment

This is an update assessment using FLICA (Kell 2007, Patterson 1998a) with the same settings as in 2008, with the 8 year separable period moved forward one year to 2001 - 2008, tuned using the complete survey time series (1987, 1991-2008). This uses catch data from 1957 to 2008 giving an assessment of $F$ from 1957 to 2008 and numbers at age from 1 Jan 1957 to 2009. The input data are given in Tables 5.6.2.1-8, the run settings are presented in Tables 5.6.2.9-11.

The results of the assessment are given as stock summary in Table 5.6.2.12 and Figure 5.6.2.1. The output values are in Tables 5.6.2.13-16. Run diagnostics are given in Tables 5.6.2.17-20 and Figures 5.6.2.2-12. The parameter estimates are given in Table 5.6.2.21.

The assessment gives an SSB for 2008 of 91884 t and a mean fishing mortality ( 3 to 6ringers) of 0.16 , the summary is given in Figure 5.6.2.1 and Table 5.6.2.12 which illustrate the stock trends from the assessment. The separable model diagnostics (Table 5.6.2.18 and Figure 5.6.2.2) show that the total residuals by age and year between the catch and separable model are reasonably trend-free. The 2000 year class is still reasonably abundant in the catch and survey data in 2008 (7-ringers). A second year class (2004, 3-ringers in 2008) is also reasonably abundant in the catch and survey data in 2008. In 2007, the catch data suggested a slightly better recruitment of the 2004 year class (2-ringers in 2007) whereas the survey suggested it was the 2003 year class (3-ringers in 2007) that was larger. It would now seem that the 2004 year class is the stronger of the two. The fits between survey and assessment are illustrated in Figures 5.6.2.3-11 for ages 1 to $9+$ winter rings. The poor fit at age 1 supports the downweighting of this index. The best fits are to middle ages 3-5.

This year's estimate of SSB for 2007 is around $92000 t$, compared with 67000 t in last year's final assessment run, an increase of $37 \%$. The assessment shows continuing low levels of recruitment (the 2001, 2002 and 2003 year classes are all weak). The tuning diagnostics (Figures 5.6.2.3 to 5.6.2.12 and Table 5.6.2.17-21) show year effects in the survey that the assessment is sensitive to. The assessment fits between negative and positive residuals in the last two years of the assessment. The analytical retrospective (Figure 5.6.2.13) plots show that the assessment is noisy but now shows a reasonably stable but historically low stock level. Although the assessment is noisy, it gives a clear indication of the state of the stock in its historical context.

The outcome of the assessment this year suggests that the SSB is relatively stable at around $15 \%$ below the average of the last 20 years, compared with the perception from last year's assessment that it was declining. Catch in 2008 is almost half the 2007 level and with the increased SSB, F has decreased to $\mathrm{F}=0.16$. Recruitment is low for the 2001, 2002 and 2003 year classes (Table 5.6.2.12). The 2004 recruitment currently appears to be around half the level of the last reasonable year class (2000); the 2005 year class appears to be around the same level as the poor 2001 - 2003 year classes. There is insufficient data to evaluate later year classes.

In conclusion, this assessment is driven by a noisy survey, giving the third lowest survey SSB estimate in 2007 to the second highest survey estimate in 2008. Point estimates of SSB and F from the survey are, therefore, not that informative and should be used to indicate medium term trends and used for guidance. The current management agreement that restricts large inter-annual changes in TACs is appropriate for such a noisy assessment.

### 5.7 Short term projections

### 5.7.1 Deterministic short-term projections

In 2005 the Working Group tested an HCR applicable to VIa (N) (ICES 2005/ACFM:16), which was accepted by ICES as precautionary. This has formed the basis for the proposed agreement and was implemented in December 2008 by the European Commission. A deterministic short-term projection is presented, which provides options including those based on the management agreement.

Short-term projections were carried out using MFDP (Smith 2000). Input data are stock numbers on $1^{\text {st }}$ January in 2009 from the 2009 ICA assessments (Section 5.6.2, Tables 5.7.1.1), with geometric mean recruitment 1989-2006 replacing recruitment both 1 - and 2-ring in 2009. For the selection of this period see productivity section in 2007 WG report. The retrospective assessment of recruitment estimates in the 2003 Working Group (ICES 2003/ACFM:17) showed the substantial revision of 1- and 2ring herring abundance ( $1^{\text {st }}$ January survivors) in subsequent assessments, justifying the use of geometric means for these ages. The selection pattern used is taken from the final year of the ICA assessment (Table 5.6.2.16, and Figure 5.6.2.2), and is therefore effectively the mean of last 8 years. For the projections, data for maturity, natural mortality, mean weights-at-age in the catch and in the stock are means of the three previous years (i.e., 2006-2008). A TAC constraint of 21760 t in 2009 is used for the basis for the intermediate year in the projection, this implies an exploitation at $\mathrm{F}=0.23$, close to target F. All the input values are summarised in Table 5.7.1.1.

The results of the short-term projection are given in Tables 5.7.1.2-5.7.1.3. For F in accordance with the proposed management plan (SSB2010 <94000t, F=0.25 in 2010 TAC increase of $12 \%$ ) catches are projected to be 24420 t , and SSB rises to approximately 96000 t in 2011.

### 5.7.2 Yield-per-recruit

Yield-per-recruit analyses were carried out using MFYPR (Smith 2000) to provide yield-per-recruit (Figure 5.7.2.1). The value for $\mathrm{F}_{0} .1$ is 0.17 .

### 5.8 Medium term projections and HCR performance \$

In 2005 ICES classed as precautionary a proposed management plan:
"An HCR with the following rule is shown to be sustainable and delivering reasonably high yield
$\mathrm{F}=0.25 \quad$ if $\mathrm{SSB}>75,000 \mathrm{t}$ Optional year on year TAC constraint.
$F=0.2$ if $S S B<75,000 t$ No constraint on TAC.
The rule should be supplemented with a requirement for $\mathrm{F}=0$ if SSB falls below Blim."

In 2008 COUNCIL REGULATION (EC) No 1300/2008 of 18 December 2008 established a multi-annual management agreement for the stock of herring distributed to the west of Scotland and the fisheries exploiting that stock.
$\mathrm{F}=0.25$ if SSB $>75000 \mathrm{t} \quad 20 \%$ TAC constraint.
$\mathrm{F}=0.20$ if SSB $<75000 \mathrm{t}$ but $>62500 \mathrm{t} \quad 20 \%$ constraint on TAC change.
$\mathrm{F}=0.20$ if SSB $<62500 \mathrm{t}$ but $>50000 \mathrm{t} \quad 25 \%$ constraint on TAC change
$\mathrm{F}=0 \quad$ if $\mathrm{SSB}<50000 \mathrm{t}$.
The agreed rule uses the same trigger points and includes the closure if SSB falls below Blim that ICES requested. However it has additional constraint on year-on-year change in TAC below 75,000 t which was not tested. In addition, ICES now provides catch options based on geometric mean recruitment from 1989 to the present (2006 this year, Section 5.7). This period was selected in 2007 following investigations by HAWG on changes in productivity of herring stocks (ICES 2007). Here we provide an exploration of the agreed rule under the new starting conditions, recruitment regimes based on SSB / recruitment 1989-2006, and typical measurement errors observed over the last few years.

### 5.8.1 Medium term simulation methods

The current investigations use the software STPR3 (Skagen 2003), the same software used to evaluate HCRs for this stock in 2005 (Simmonds and Keltz 2007), and used by WKHMP in 2008 (ICES 2008/ACOM:27). Parameterisation follows the principles used in Simmonds and Keltz (2007) and in the Stock Annex (06), with values updated to account for new data in the assessment input and estimates of SSB and recruits. These are as follows:

- Numbers at age 1 of January 2009 from input to short term forecast (Table 5.7.1.1)
- Precision of starting numbers from the covariance matrix from ICA, except for the variance for 1-ring which is not properly estimated so it was set equal to the variance for 2 -ring $=0.117$ (equivalent to a CV ~35\%) Table 5.8.1.1
- Mean weights for the catch (Table 5.6.2.2) and stock (Table 5.6.2.3) and maturities (Table 5.6.2.5) from 1990 to 2008 were selected randomly as year sets within the simulations. Prior to 1990 the mean weights and maturities for the stock were not well estimated.
- Natural mortality (Table 5.6.2.4) and selection (Table 5.6.2.16) matching the assessment.
- Catch in tonnes without bias (catch is usually not fully taken) with $5 \% \mathrm{CV}$ to mimic small uncertainty. Options of positive bias (underreporting) were tested to explore the robustness of the rule.
- Assessment error two options tested
- $25 \%$ derived from ICES historic assessment database 2000 to 2008.
- $30 \%$ for robustness testing.
- Stock / Recruit relationship two options (Table 5.8.1.2)
- Beverton-Holt model fitted to S/R pairs (Table 5.6.2.12, shifted by 2 years) using FLR (Kell 2007) for years 1989-2006 using slope from fit to 1957 to 2006 (Figure 5.8.1.1), as slope on 1989-2006 is unresolved (Figure 5.8.1.2).
- Fitted Hockey-Stick (segmented regression) to S/R pairs (Table 5.6.2.12, shifted by 2 years) for years 1989-2006 with breakpoint at lowest observed biomass at 50000 t as no breakpoint is found within the observations.
- Stochastic draws lognormal truncated at 1.5 to give correct proportions of numbers of observations in the tails of the probability distribution (Figure 5.8.1.3).
- Recruitment year-year autocorrelation +0.16 (as observed, though not significant)


### 5.8.2 Medium term simulation results

The results for two S/R model options (run 5 - Hockey-Stick and run 1 Beverton-Holt) are presented in Figure 5.8.2.1. These show that in both cases initial risks are around $8 \%$. These reduce over time to below $5 \%$ and continue to reduce to less than $1 \%$ as the SSB increases. These initial risks result from the starting conditions, with SSB below equilibrium and variability which can be greater than that in the future due to uncertainty in these numbers from the ICA assessment. The run using the Hockey-Stick S/R assumption assumes slightly higher recruitment at lower stock size, giving faster stock growth, and more rapidly reducing risk of $\mathrm{SSB}<\mathrm{Blim}$. The Beverton-Holt $\mathrm{S} / \mathrm{R}$ assumption allows recruitment to continue to grow with SSB and results in a higher biomass after a number of years and lower risk in the longer term as SSB continues to increase beyond the end of the 10 years illustrated.

In order to explore the robustness of these evaluations we have compared risks of $\mathrm{SSB}<\mathrm{Blim}$ under a range of different assumptions (Figure 5.8.2.2).

In the medium term all risks decline below $5 \%$ and decline to very low levels with the exception of those with positive bias in the catch of 20 and $30 \%$ (Runs 10-11).

Results of runs with the two types of year-on-year constraint on TAC (Figure 5.8.2.2a, dotted $20 \%$ y-y constraint (runs 1, 3, 5, 7), solid $25 \%$ y-y constraint (runs $2,4,6,8$ )), indicate that risk of $\mathrm{SSB}<$ Blim with either of these constraints or the current rule with change at 62500 t are effectively the same.

Risks are higher if the measurement error is increased from a CV of $25 \%$ (Runs (1, 2, 5, 6)) to 30\% (Runs (3, 4, 7, 8)). (Figure 5.8.2.2a)

The risk increases when unregulated catch is assumed to increases from 0 to 10, 20 and $30 \%$ (Runs 9, 10, 11 respectively) (Figure 5.8.2.2b). Only at the highest level of unregulated catch ( $30 \%$ ) never seen in this area do the risks increase to levels approaching 3\% in the long term.

In conclusion, the current state of the stock implies about $8 \%$ risk of $\mathrm{SSB}<\mathrm{Blim}$. Following the agreed management plan the risk is expected to decrease to well below the precautionary limit reference of $5 \%$. The time over which this reduction in risk will occur is dependent on the realised recruitment, though the most plausible assumptions show a rapid decline. The changes to the previously ICES endorsed precautionary management plan are small. Changes to recruitment increase the risks slightly but these are expected to be well below $5 \%$. Therefore it is recommended that the amended plan be accepted as precautionary.

### 5.9 Precautionary and yield based reference points

Blim is agreed at $50000 t$ (based on Bloss). There are no agreed precautionary reference points for this stock. The agreed management rule has a Btrig at 75000 t .

### 5.10 Quality of the Assessment

The HAWG considers that this year's assessment is as reliable as last year's. The precision of the assessment estimated through parametric bootstrap is shown in Figure 5.10.1. The influence of model settings has been explored and shown to give some differences but does not change the conclusions that $F$ is below target F and SSB is above $\mathrm{B}_{\text {pa }}$. The assessment outcomes were revised upwards from those made last year. SSB, catch and F estimated in last year's assessment and short term forecast are compared with this year's assessment in the text table below.

|  | 2008 REPORT |  |  |  |  | This year |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Year | SSB | Catch | F 3-6 |  | Year | SSB | Catch | F 3-6 |
| Assess | 2006 | 76813 | 27346 | 0.28 | Assess | 2006 | 93270 | 27346 | 0.23 |
| 2008 | 2007 | 68816 | 29616 | 0.40 | 2009 | 2007 | 91848 | 29616 | 0.29 |
| STF* 2008 | 2008 | 68 444* | 13 011* | 0.20* |  | 2008 | 91884 | 16054 | 0.16 |

* projected values from the intermediate year in the deterministic short term forecast assuming a catch consistent with management plan. STF refers to values estimated in the first year of the short term forecast in the 2008 report.
Retrospective analyses of the assessment from 2008 to 2004 (Figure 5.6.2.13) support the perception of a noisy but fairly well balanced assessment. Catches are below TACs; recruitment is low.


### 5.11 Management Considerations

In the absence of precautionary reference points the state of the stock cannot be evaluated. An analytical assessment shows that SSB (in 2009) is 1.8 times Blim. ICES considers that the stock is currently fluctuating at a low level and is being exploited close to Fmsy. Recruitment has been low since 1998, and the 2001, 2002 and 2003 year classes are weak.

There has been considerable uncertainty in the amount of landings from this stock in the past. Area misreporting continues to be a problem, with almost all countries taking catches of herring in other areas and reporting it into VIa (N). Increased observer coverage and or use of VMS and electronic log books might reduce these problems.

The assessment is noisy, leading to annual revisions of SSB and F. The management plan has been designed to cope with this by applying a constraint on year-on-year change in TAC. Revisions in SSB can be upwards or downwards, so it is important to maintain the restrictions on change in TAC both when the stock is revised upwards or downwards. Asymmetrical changes in TAC have not been tested and may be significantly more risky than those tested.

The stock identity of herring west of the British Isles was reviewed by the EU-funded project WESTHER. This identified Division VIa (N) as an area where catches comprise a mixture of fish from Divisions VIa (N), VIa (S), and VIIa (N). Concerning the management plan for Division VIa (N), ICES has advised that herring components should be managed separately to afford maximum protection. If there is an increasing catch on the mixed fishery in Division VIa (N), this should be considered in the management of the Division VIa (S) component which is in a depleted state. In 2008 ICES has began to evaluate management for this Division VIa (S) and VIIa (N). It will be a
number of years before ICES can provide a fully operational integrated strategy for these units. In this context ICES recommends that the management plan for Division VIa (N) should be continued.

### 5.12 Ecosystem Considerations

Herring are an important prey species in the ecosystem and also one of the dominant planktivorous fish.

Observers monitor the fisheries. Herring fisheries tend to be clean with little bycatch of other fish. Scottish discard observer programs since 1999 and more recently Dutch observers indicate that discarding of herring in these directed fisheries is at a low level. The Scottish discard observer programs have recorded occasional catches of seals and zero catches of cetaceans.

### 5.13 Changes in the environment

Temperatures in this area have been increasing over the last number of decades. There are indications that salinity is also increasing (ICES 2006). It is considered that this may have implications for herring. It is known that similar environmental changes have affected the North Sea herring. There is evidence that there have been recent changes of the productivity of this stock (ICES HAWG 2007).

Herring are thought to be a source of food for seals. Grey seals (Halichoerus grypus) are common in many parts of the Celtic Seas area. The majority of individuals are found in the Hebrides and in Orkney (SCOS 2005). A recent study (Hammond \& Harris 2006) of seal diets off western Scotland revealed that grey seals may be an important predator for cod, herring and sandeels in this area. Common seals (Phoca vitulina) are also widespread in the northern part of the ecoregion with around 15,000 animals estimated (SCOS 2005). The numbers of seals in VIa (N) is thought to have increased over the last decades. The seal consumption of herring is estimated with great uncertainty and the impact of increased predation is not known, but there is a possibility that seal predation could influence natural mortality.

Table 5.1.1. Herring in VIa (N). Catch in tonnes by country, 1985-2008. These figures do not in all cases correspond to the official statistics and cannot be used for management purposes.

| Country | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  |  |
| Faroes | 104 | 400 |  |  |  | 326 | 482 |  |
| France | 20 | 18 | 136 | 44 | 1342 | 1287 | 1168 | 119 |
| Germany | 5937 | 2188 | 1711 | 1860 | 4290 | 7096 | 6450 | 5640 |
| Ireland |  | 6000 | 6800 | 6740 | 8000 | 10000 | 8000 | 7985 |
| Netherlands | 5500 | 5160 | 5212 | 6131 | 5860 | 7693 | 7979 | 8000 |
| Norway | 4690 | 4799 | 4300 | 456 |  | 1607 | 3318 | 2389 |
| UK | 28065 | 25294 | 26810 | 26894 | 29874 | 38253 | 32628 | 32730 |
| Unallocated | -502 | 37840 | 18038 | 5229 | 2123 | 2397 | -10597 | -5485 |
| Discards |  |  |  |  | 1550 | 1300 | 1180 | 200 |
| Total | 43814 | 81699 | 63007 | 47354 | 53039 | 69959 | 50608 | 51578 |
| Area-Misreported | -4672 | -10935 | -18647 | -11763 | -19013 | -25266 | -22079 | -22593 |
| WG Estimate | 39142 | 70764 | 44360 | 35591 | 34026 | 44693 | 28529 | 28985 |
| Source (WG) | 1987 | 1988 | 1989 | 1990 | 1991 | 1993 | 1993 | 1994 |
| Country | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| Faroes |  |  |  |  |  |  |  |  |
| France | 818 | 274 | 3672 | 2297 | 3093 | 1903 | 463 | 870 |
| Germany | 4693 | 5087 | 3733 | 7836 | 8873 | 8253 | 6752 | 4615 |
| Ireland | 8236 | 7938 | 3548 | 9721 | 1875 | 11199 | 7915 | 4841 |
| Netherlands | 6132 | 6093 | 7808 | 9396 | 9873 | 8483 | 7244 | 4647 |
| Norway | 7447 | 8183 | 4840 | 6223 | 4962 | 5317 | 2695 |  |
| UK | 32602 | 30676 | 42661 | 46639 | 44273 | 42302 | 36446 | 22816 |
| Unallocated | -3753 | -4287 | -4541 | -17753 | -8015 | -11748 | -8155 |  |
| Discards |  | 700 |  |  | 62 | 90 |  |  |
| Total | 56175 | 54664 | 61271 | 64359 | 64995 | 65799 | 61514 | 37789 |
| Area-Misreported | -24397 | -30234 | -32146 | -38254 | -29766 | -32446 | -23623 | -19467 |
| WG Estimate | 31778 | 24430 | 29575 | 26105 | 35233* | 33353 | 29736 | $18322{ }^{\text {s }}$ |
| Source (WG) | 1995 | 1996 | 1997 | 1997 | 1998 | 1999 | 2000 | 2001 |
| Country | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| Faroes |  | 800 | 400 | 228 | 1810 | 570 | 484 | 927 |
| France | 760 | 1340 | 1370 | 625 | 613 | 701 | 703 | 564 |
| Germany | 3944 | 3810 | 2935 | 1046 | 2691 | 3152 | 1749 | 2526 |
| Ireland | 4311 | 4239 | 3581 | 1894 | 2880 | 4352 | 5129 | 3103 |
| Netherlands | 4534 | 4612 | 3609 | 8232 | 5132 | 7008 | 8052 | 4133 |
| Norway |  |  |  |  |  |  |  |  |
| UK | 21862 | 20604 | 16947 | 17706 | 17494 | 18284 | 17618 | 13963 |
| Unallocated |  | 878 | -7 |  |  |  |  |  |
| Discards |  |  |  | 123 | 772 | 163 |  |  |
| Total | 35411 | 36283 | 28835 | 29854 | 31392 | 34230 | 33735 | 25216 |
| Area-Misreported | -11132 | -8735 | -3581 | -7218 | -17263 | -6884 | -4119 | -9162 |
| WG Estimate | $24556^{\text {\$ }}$ | 32914 ${ }^{\text {8 }}$ | 28081\$ | $25021^{\text {\$ }}$ | 14129 ${ }^{\text { }}$ | 27346 | 29616 | 16054 |
| Source (WG) | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |

\$Revised at HAWG 2007

Table 5.2.1. Herring in VIa (N). Catch and sampling effort by nations participating in the fishery in 2008.

| PERIOD : 1 |  |  |
| :---: | ---: | ---: |
| Country | Sampled <br> Catch | Official <br> Caroes |
| Freland | 0.00 | 517.00 |
| Netherlands | 0.00 | 1337.00 |
| Scotland | 0.00 | 278.00 |
| Period Total | 0.00 | 132.00 |
| Sum of Offical Catches : | 0.00 | 2264.00 |
| Unallocated Catch : | 2264.00 |  |
| Working Group Catch : | -1615.00 |  |
|  |  | 649.00 |

PERIOD : 2


PERIOD : 4

| Country | Sampled <br> Catch | Official <br> Catch |
| :--- | ---: | ---: |
| England \& Wales | 0.00 | 1859.00 |
| Faroes | 0.00 | 25.00 |
| Ireland | 0.00 | 1766.00 |
| N. Ireland | 0.00 | 204.00 |
| Scotland | 0.00 | 5.00 |
| Period Total | 0.00 | 3859.00 |
| Sum of Offical Catches : | 3859.00 |  |
| $\quad$ Unallocated Catch : |  | -1766.00 |
| Working Group Catch : | 2093.00 |  |


| No. of | No. | No. | SOP |
| :---: | :---: | ---: | :---: |
| samples | measured | aged | $\%$ |
| 0 | 0 | 0 | 0.00 |
| 0 | 0 | 0 | 0.00 |
| 0 | 0 | 0 | 0.00 |
| 0 | 0 | 0 | 0.00 |
| 0 | 0 | 0 | 0.00 |
| 0 | 0 | 0 | 0.00 |

Total over all Areas and Periods

| Country | Sampled Catch | Official Catch |
| :---: | :---: | :---: |
| England \& Wales | 1811.00 | 3670.00 |
| Faroes | 0.00 | 927.00 |
| France | 0.00 | 564.00 |
| Germany | 0.00 | 2526.00 |
| Ireland | 0.00 | 3103.00 |
| N. Ireland | 0.00 | 204.00 |
| Netherlands | 964.00 | 4133.00 |
| Scotland | 7062.00 | 10089.00 |
| Total for Stock | 9837.00 | 25216.00 |
| Sum of Offical Catches : |  | 25216.00 |
| Unallocated Catch : |  | -9162.00 |
| Working Group Catch : |  | 16054.00 |


| No. of <br> samples | No. <br> measured | No. | aged |
| :---: | :---: | ---: | :---: |

Table 5.3.1. Herring in VIa (N). Estimates of abundance, biomass, maturity, weight- and length-atage from Scottish acoustic surveys. Thousands of fish at age and spawning biomass (SSB, tonnes). N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Age ( ring) | Numbers | Biomass | Maturity | weight(g) | Length (cm) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |
| 1 | 47.84 | 2.6 | 0.00 | 54.6 | 18.2 |
| 2 | 232.57 | 40.0 | 0.98 | 172.1 | 26.3 |
| 3 | 911.95 | 174.5 | 1.00 | 191.3 | 27.2 |
| 4 | 668.87 | 139.3 | 1.00 | 208.3 | 28.0 |
| 5 | 339.92 | 72.8 | 1.00 | 214.3 | 28.2 |
| 6 | 272.23 | 58.2 | 1.00 | 213.9 | 28.2 |
| 7 | 720.86 | 159.0 | 1.00 | 220.6 | 28.5 |
| 8 | 365.89 | 82.0 | 1.00 | 224.2 | 28.6 |
| $9+$ | 263.74 | 62.9 | 1.00 | 238.5 | 29.2 |
| Immature | 53.461 | 3.3 |  | 61.2 | 18.7 |
| Mature | 3770.421 | 788.1 |  | 209.0 | 28.0 |
| Total | 3823.882 | 791.4 | 0.99 | 207.0 | 27.9 |

Tables 5.6.2.1. - 5.6.2.21. Herring in VIa (N). Input data, FLICA run settings and results for the maximum-likelihood ICA calculation for the 8 year separable period. N.B. In these tables "age" refers to number of rings (winter rings in the otolith).

TABLE 5.6.2.1 HERRING in VIa (N). CATCH IN NUMBER
Units : Thousands

table 5.6.2.2 herring in VIa (N). WEIGHTS AT AGE IN THE CATCH

```
Units : Kg
    year
age 1957 1958 1959 1960 1961 196 1962 1963 1964 1965 1966 1967 1968
    1 0.079 0.079 0.079 0.079 0.079 0.079 0.079 0.079 0.079 0.079 0.079 0.079
    2 0.104 0.104 0.104 0.104 0.104 0.104 0.104 0.104 0.104 0.104 0.104 0.104
    3 0.130 0.130 0.130 0.130 0.130 0.130 0.130 0.130 0.130 0.130 0.130 0.130
    4 0.158 0.158 0.158 0.158 0.158 0.158 0.158 0.158 0.158 0.158 0.158 0.158
    5 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164
    6 0.170 0.170 0.170 0.170 0.170 0.170 0.170 0.170 0.170 0.170 0.170 0.170
    7 0.180 0.180 0.180 0.180 0.180 0.180 0.180 0.180 0.180 0.180 0.180 0.180
    8 0.183 0.183 0.183 0.183 0.183 0.183 0.183 0.183 0.183 0.183 0.183 0.183
    9 0.185 0.185 0.185 0.185 0.185 0.185 0.185 0.185 0.185 0.185 0.185 0.185
        year
age 1969 1970 1971 1972 1973 1974 1975 1976 1977 1978 1979 1980
    1 0.079 0.079 0.079 0.079 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090
    2 0.104 0.104 0.104 0.104 0.121 0.121 0.121 0.121 0.121 0.121 0.121 0.121
    3 0.130 0.130 0.130 0.130 0.158 0.158 0.158 0.158 0.158 0.158 0.158 0.158
    4 0.158 0.158 0.158 0.158 0.175 0.175 0.175 0.175 0.175 0.175 0.175 0.175
    5 0.164 0.164 0.164 0.164 0.186 0.186 0.186 0.186 0.186 0.186 0.186 0.186
    6 0.170 0.170 0.170 0.170 0.206 0.206 0.206 0.206 0.206 0.206 0.206 0.206
    7 0.180 0.180 0.180 0.180 0.218 0.218 0.218 0.218 0.218 0.218 0.218 0.218
    8 0.183 0.183 0.183 0.183 0.224 0.224 0.224 0.224 0.224 0.224 0.224 0.224
    9 0.185 0.185 0.185 0.185 0.224 0.224 0.224 0.224 0.224 0.224 0.000 0.000
        year
age 1981 1982 1983 1984 1985
    1 0.090 0.080 0.080 0.080 0.069 0.113 0.073 0.080 0.082 0.079 0.084 0.091
    2 0.121 0.140 0.140 0.140 0.103 0.145 0.143 0.112 0.142 0.129 0.118 0.119
    3 0.158 0.175 0.175 0.175 0.134 0.173 0.183 0.157 0.145 0.173 0.160 0.183
    4 0.175 0.205 0.205 0.205 0.161 0.196 0.211 0.177 0.191 0.182 0.203 0.196
    5 0.186 0.231 0.231 0.231 0.182 0.215 0.220 0.203 0.190 0.209 0.211 0.227
    6 0.206 0.253 0.253 0.253 0.199 0.230 0.238 0.194 0.213 0.224 0.229 0.219
    7 0.218 0.270 0.270 0.270 0.213 0.242 0.241 0.240 0.216 0.228 0.236 0.244
    8 0.224 0.284 0.284 0.284 0.223 0.251 0.253 0.213 0.204 0.237 0.261 0. 256
    9 0.224 0.295 0.295 0.295 0.231 0.258 0.256 0.228 0.243 0.247 0.271 0. 256
        year
age 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004
    1 0.089 0.083 0.106 0.081 0.089 0.097 0.076 0.083 0.049 0.107 0.060 NaN
    2 0.128 0.142 0.142 0.134 0.136 0.138 0.130 0.137 0.140 0.146 0.145 0.154
    3 0.158 0.167 0.181 0.178 0.177 0.159 0.158 0.164 0.163 0.163 0.160 0.173
    4 0.197 0.190 0.191 0.210 0.205 0.182 0.175 0.183 0.183 0.173 0.169 0.195
    5 0.206 0.195 0.198 0.230 0.222 0.199 0.191 0.201 0.192 0.160 0.186 0.216
    6 0.228 0.201 0.214 0.233 0.223 0.218 0.210 0.215 0.196 0.179 0.200 0.220
    7 0.223 0.244 0.208 0.262 0.219 0.227 0.225 0.239 0.205 0.187 0.194 0.199
    8 0.262 0.234 0.227 0.247 0.238 0.212 0.223 0.281 0.225 0.245 0.186 0.190
    9 0.263 0. 266 0.277 0.291 0.263 0.199 0.226 0.253 0.272 0.281 0.294 0.311
        year
age 2005 2006 2007 2008
    1 0.108 0.091 0.115 NaN
    2 0.133 0.158 0.167 0.170
    3 0.163 0.168 0.188 0. 206
    4 0.184 0.193 0.197 0.231
    5 0.211 0.208 0.210 0.231
    6 0.226 0.225 0.221 0.249
    7 0.234 0.244 0.216 0.253
    8 0.256 0.262 0.262 0.284
    9 0.250 0.275 0.303 0.288
```

table 5.6.2.3 herring in VIa ( N ). WEIGHTS AT AGE IN the stock

```
Units : Kg
    year
age 1957 1958 1959 1960 1961 196 1962 1963 1964 1965 1966 1967 1968
    1 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090
    2 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164
    3 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0. 208
    4 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233
    5 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246
    6 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0. 252
    7 0.258 0. 258 0. 258 0.258 0.258 0. 258 0. 258 0. 258 0. 258 0. 258 0. 258 0. 258
    8 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269
    9 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0. 292
        year
age 1969 1970 1971 1972 1973 1974 1975 1976 1977 1978 1979 1980
    1 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090
    2 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164
    3 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208
    4 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0. 233
    5 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246
    6 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0.252 0.252
    7 0.258 0. 258 0. 258 0.258 0. 258 0. 258 0. 258 0. 258 0. 258 0. 258 0. 258 0. 258
    8 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269
    0 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0. 292 0.000 0.000
        year
age 1981 1982 1983 1984 1985
    1 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090 0.090
    2 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164 0.164
    3 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208 0.208
    4 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0.233 0. 233
    5 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246 0.246
    6 0.252 0. 252 0. 252 0.252 0.252 0.252 0.252 0. 252 0. 252 0.252 0. 252 0. 252
    7 0.258 0.258 0.258 0.258 0.258 0.258 0.258 0.258 0.258 0.258 0.258 0.258
    8 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269 0.269
    9 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0.292 0. 292 0.292 0. 292
        year
age 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004
    1 0.075 0.052 0.042 0.045 0.057 0.066 0.054 0.062 0.062 0.062 0.064 0.059
    2 0.162 0.150 0.144 0.140 0.150 0.138 0.137 0.141 0.132 0.153 0.138 0.138
    3 0.196 0.192 0.191 0.180 0.189 0.176 0.166 0.173 0.170 0.177 0.176 0.159
    4 0.206 0.220 0.202 0.209 0.209 0.194 0.188 0.183 0.190 0.198 0.190 0.180
    5 0.226 0.221 0.225 0.219 0.225 0.214 0.203 0.194 0.198 0.212 0.204 0.189
    6 0.234 0.233 0.227 0.222 0.233 0.226 0.219 0.204 0.212 0.215 0.213 0.202
    7 0.254 0.241 0.247 0.229 0.248 0.234 0.225 0.211 0.220 0.225 0.217 0.213
    8 0.260 0.270 0.260 0.242 0.266 0.225 0.235 0.222 0.236 0.243 0.223 0.214
    9 0.276 0.296 0.293 0.263 0.287 0.249 0.245 0.230 0.254 0.259 0.228 0. 206
        year
age 2005 2006 2007 2008
    1 0.075 0.075 0.075 0.055
    2 0.130 0.135 0.168 0.172
    30.154 0.166 0.183 0.191
    4 0.166 0.185 0.191 0.208
    5 0.180 0.192 0.195 0.214
    6 0.191 0.204 0.195 0.214
    7 0.212 0.211 0.202 0.221
    0.203 0.224 0.203 0.224
    0 0.228 0.231 0.214 0.239
```

TABLE 5.6.2.4 HERRING in VIa (N). NATURAL MORTALITY

| Un | $\begin{aligned} & \text { ts : } \\ & \text { year } \end{aligned}$ | NA |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1957 | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 |
| 1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| 2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |
| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 5 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 7 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 8 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 9 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| 2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |
| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 5 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 7 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 8 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 9 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| 2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |
| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 5 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 7 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 8 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 9 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |  |  |  |  |  |  |  |  |
| 1 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |  |  |  |  |  |  |  |  |
| 2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |  |  |  |  |  |  |  |  |
| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |  |  |  |  |  |  |  |  |
| 4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |
| 5 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |
| 6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |
| 7 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |
| 8 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |
| 9 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |

TABLE 5.6.2.5 HERRING in VIa (N). PROPORTION MATURE

```
Units : NA
    year
age 1957 1958 1959 1960 1961 1962 1963 1964 1965 1966 1967 1968 1969 1970 1971
    1 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00
    2 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57
    3 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96
    4 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    5 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    6 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    7 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    8 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    9 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    year
age 1972 1973 1974 1975 1976 1977 1978 1979 1980 1981 1982 1983 1984 1985 1986
    1 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00
    2 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57 0.57
    3 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96 0.96
    4 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    5 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    6 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    7 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    8 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    91.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
        year
age 1987 1988 1989 1990 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001
    1 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00
    2 0.57 0.57 0.57 0.57 0.57 0.47 0.93 0.48 0.19 0.76 0.57 0.85 0.57 0.45 0.93
    3 0.96 0.96 0.96 0.96 0.96 1.00 0.96 0.92 0.98 0.94 0.96 0.97 0.98 0.92 0.99
    4 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    5 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    6 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    7 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    8 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    91.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00 1.00
        year
age 2002 2003 2004 2005 2006 2007 2008
    1 0.00 0.00 0.00 0.00 0.00 0 0.00
    2 0.92 0.76 0.83 0.84 0.81 1 0.98
    31.00 1.00 0.97 1.00 0.96 1 1.00
    4 1.00 1.00 1.00 1.00 1.00 1 1.00
    5 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    6 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    71.00 1.00 1.00 1.00 1.00 1.00 1.00
    8 1.00 1.00 1.00 1.00 1.00 1.00 1.00
    91.00 1.00 1.00 1.00 1.00 1 1.00
```

TABLE 5.6.2.6 HERRING in VIa (N). FRACTION OF HARVEST BEFORE SPAWNING

```
Units : NA
    year
age 1957 1958 1959 1960 1961 1962 1963 1964 1965 1966 1967 1968 1969 1970 1971
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
        year
age 1972 1973 1974 1975 1976 1977 1978 1979 1980 1981 1982 1983 1984 1985 1986
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
        year
age 1987 1988 1989 1990 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
        year
age 2002 2003 2004 2005 2006 2007 2008
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67
```

table 5.6.2.7 HERRING in VIa (N). FRACTION OF NATURAL MORTALITY BEFORE SPAWNING

```
Units : NA
    year
age 1957 1958 1959 1960 1961 1962 1963 1964 1965 1966 1967 1968 1969 1970 1971
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
        year
age 1972 1973 1974 1975 1976 1977 1978 1979 1980 1981 1982 1983 1984 1985 1986
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
        year
age 1987 1988 1989 1990 1991 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67 0.67
        year
age 2002 2003 2004 2005 2006 2007 2008
    1 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    2 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    3 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    4 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    5 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    6 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    7 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    8 0.67 0.67 0.67 0.67 0.67 0.67 0.67
    9 0.67 0.67 0.67 0.67 0.67 0.67 0.67
```

TABLE 5.6.2.8 HERRING in VIa (N). SURVEY INDICES
FLT01:West Scotland Summer Acoustic Survey (Catch:Thousands)(Effort:Unknown) Configuration


FLT01:West Scotland Summer Acoustic Survey (Catch:Thousands)(Effort:Unknown) Index Values

| $\begin{aligned} & \text { ts : NA } \\ & \text { year } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1987 | 19881989 | 1990 | 1991 | 1992 | 19931 | 41995 |  | 1996 | 1997 |
| 1 | 249100 | -1 -1 | -1 | 338312 | 74310 | 27604941 | 4606304 |  | 2207923 | 792320 |
| 2 | 578400 | -1 -1 | -1 | 2944845 | 503430750 | 52705420 | 0108509057 |  | 460641 | 641860 |
| 3 | 551100 | -1 -1 | -1 | 327902 | 210980681 | 11706077 | 47271080 |  | 2530281 | 286170 |
| 4 | 353100 | -1 -1 | -1 | 367830 | 258090653 | 30502856 | 45025032 |  | 110167 | 167040 |
| 5 | 752600 | -1 -1 | -1 | 488288 | 414750544 | 40003067 | 153000 |  | 66100 |  |
| 6 | 111600 | -1 -1 | -1 | 176348 | 240110865 | 1502681 | 187060 |  | 60049 | 49520 |
| 7 | 48100 | -1 -1 | -1 | 98741 | 105670284 | 41104068 | 169180 |  | 380162 | 16280 |
| 8 | 15900 | -1 -1 | -1 | 89830 | 567101517 | 17301737 | 2365807 |  | 19028 | 28990 |
| 9 | 6500 | -1 -1 | -1 | 58043 | 634401561 | 1801318 | 201 | 510114 | 81024 | 440 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1998 | 1999 | 2000 | 2001 | 12002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 1 | 1221700 | 534200 | 447600 | 313100 | 424700 | 438800 | 564000 | 50200 | 112300 | -1 |
| 2 | 794630 | 322400 | 316200 | 1062000 | 436000 | 1039400 | 274500 | 243400 | 835200 | 126000 |
| 3 | 666780 | 1388000 | 337100 | 217700 | 1436900 | 932500 | 760200 | 230300 | 387900 | 294400 |
| 4 | 471070 | 432000 | 899500 | 172800 | 0199800 | 1471800 | 442300 | 423100 | 284500 | 202500 |
| 5 | 179050 | 308000 | 393400 | 437500 | - 161700 | 181300 | 577200 | 245100 | 582200 | 145300 |
| 6 | 79270 | 138700 | 247600 | 132600 | - 424300 | 129200 | 55700 | 152800 | 414700 | 346900 |
| 7 | 28050 | - 86500 | 199500 | 102800 | 152300 | 346700 | 61800 | 12600 | 227000 | 242900 |
| 8 | 13850 | - 27600 | 95000 | 52400 | 07500 | 114300 | 82200 | 39000 | 21700 | 163500 |
| 9 | 36770 | - 35400 | 65000 | 34700 | 059500 | 75200 | 76300 | 26800 | 59300 | 32100 |
| year |  |  |  |  |  |  |  |  |  |  |
| age 2008 | 2008 |  |  |  |  |  |  |  |  |  |
| 147840 |  |  |  |  |  |  |  |  |  |  |
| 2232570 |  |  |  |  |  |  |  |  |  |  |
| 3 | 911950 |  |  |  |  |  |  |  |  |  |
| 4668870 |  |  |  |  |  |  |  |  |  |  |
| 5 | 339920 |  |  |  |  |  |  |  |  |  |
| 6272230 |  |  |  |  |  |  |  |  |  |  |
| 7 | 720860 |  |  |  |  |  |  |  |  |  |
| 8365890 |  |  |  |  |  |  |  |  |  |  |
| 9 | 263740 |  |  |  |  |  |  |  |  |  |

FLT01:West Scotland Summer Acoustic Survey (Catch:Thousands)(Effort:Unknown) Index Variance (Inverse Weights)

| Uni | $\begin{aligned} & \text { ts : } \\ & \text { year } \end{aligned}$ | NA |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 |
| 2 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 3 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 5 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

age 2002200320042005200620072008

| 1 | 10 | 10 | 10 | 10 | 10 | 10 | 10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 3 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 5 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

TABLE 5.6.2.9 HERRING in VIa (N). STOCK OBJECT CONFIGURATION

| min | max plusgroup | minyear | maxyear | minfbar | maxfbar |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 9 | 9 | 1957 | 2008 | 3 |

TABLE 5.6.2.10 HERRING in VIa (N). FLICA CONFIGURATION SETTINGS

| sep. 2 | : NA |
| :---: | :---: |
| sep.gradual | TRUE |
| sr | : FALSE |
| sr.age | : 1 |
| lambda.age | : 0.11111111110 |
| lambda.yr | : 111111111 |
| lambda.sr | : 0.01 |
| index.model | : linear |
| index.cor | 1 |
| sep.nyr | : 8 |
| sep.age | : 4 |
| sep.sel | 1 |

TABLE 5.6.2.11 HERRING in VIa (N). FLR, R SOFTWARE VERSIONS
R version 2.8.1 (2008-12-22)

```
Package : FLICA
Version : 1.4-10
Packaged : Sat Mar 21 18:30:56 2009; mpa
Built : R 2.8.0; ; 2009-03-21 18:30:58; windows
Package : FLAssess
Version : 1.99-102
Packaged : Sun Mar 22 12:18:48 2009; mpa
Built : R 2.8.0; i386-pc-mingw32; 2009-03-22 12:18:51; windows
Package : FLCore
Version : 3.0
Packaged : Tue Mar 10 04:42:26 2009; theussl
Built : R 2.8.1; i386-pc-mingw32; 2009-03-10 04:42:28; windows
```

TABLE 5.6.2.12 HERRING in VIa (N). STOCK SUMMARY

| Year | Recruitment | TSB | SSB | Fbar | Landings <br> Tonnes | Landings SOP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | e 1 |  |  | (Ages 3-6) |  |  |
|  |  |  |  | f |  |  |
| 1957 | 1085415 | 405341 | 184542 | 0.2832 | 43438 | 0.7258 |
| 1958 | 2129925 | 498370 | 200966 | 0.3315 | 59669 | 0.7470 |
| 1959 | 2124103 | 533658 | 214361 | 0.3042 | 65221 | 0.7248 |
| 1960 | 628969 | 428836 | 248252 | 0.1948 | 63759 | 0.5679 |
| 1961 | 1282671 | 435629 | 248296 | 0.1290 | 46353 | 0.5846 |
| 1962 | 2323456 | 543692 | 237651 | 0.2055 | 58195 | 0.7727 |
| 1963 | 2128325 | 576715 | 261562 | 0.1830 | 49030 | 0.6970 |
| 1964 | 979318 | 526206 | 307678 | 0.1530 | 64234 | 0.5774 |
| 1965 | 7855652 | 1121416 | 316322 | 0.1580 | 68669 | 0.8586 |
| 1966 | 1065520 | 850597 | 427781 | 0.1920 | 100619 | 1.0136 |
| 1967 | 2499919 | 833400 | 460184 | 0.1885 | 90400 | 0.8072 |
| 1968 | 4100323 | 955297 | 437487 | 0.1425 | 84614 | 0.7964 |
| 1969 | 2998830 | 984297 | 475825 | 0.2415 | 107170 | 0.7573 |
| 1970 | 3440170 | 1002417 | 444140 | 0.3580 | 165930 | 0.7343 |
| 1971 | 9572399 | 1515913 | 316111 | 0.7885 | 207167 | 1.0162 |
| 1972 | 2675839 | 1116445 | 443971 | 0.3648 | 164756 | 1.0239 |
| 1973 | 1074339 | 801989 | 385359 | 0.6055 | 210270 | 1.0438 |
| 1974 | 1672283 | 576517 | 204084 | 0.9570 | 178160 | 1.1255 |
| 1975 | 2101533 | 434638 | 107141 | 0.9092 | 114001 | 1.0108 |
| 1976 | 608221 | 263752 | 73429 | 1.0677 | 93642 | 0.9984 |
| 1977 | 621969 | 162995 | 51907 | 0.9935 | 41341 | 0.9154 |
| 1978 | 913517 | 170722 | 48526 | 0.6768 | 22156 | 1.0056 |
| 1979 | 1216369 | 215915 | 72378 | 0.0008 | 60 | 1.0011 |
| 1980 | 885405 | 252213 | 122146 | 0.0002 | 306 | 1.0007 |
| 1981 | 1660598 | 364460 | 131858 | 0.3622 | 51420 | 0.9698 |
| 1982 | 770261 | 305592 | 109542 | 0.6750 | 92360 | 1.0347 |
| 1983 | 2977418 | 426873 | 81150 | 0.7138 | 63523 | 1.0277 |
| 1984 | 1132203 | 353359 | 120051 | 0.5182 | 56012 | 0.9494 |
| 1985 | 1199475 | 348549 | 147680 | 0.3157 | 39142 | 1.0058 |
| 1986 | 887707 | 314089 | 133119 | 0.5272 | 70764 | 1.0479 |
| 1987 | 2097344 | 380177 | 123340 | 0.3442 | 44360 | 0.9725 |
| 1988 | 899294 | 334214 | 147822 | 0.2858 | 35591 | 1.0236 |
| 1989 | 844942 | 318192 | 163947 | 0.2480 | 34026 | 1.0199 |
| 1990 | 433443 | 269972 | 154485 | 0.3500 | 44693 | 0.9889 |
| 1991 | 380590 | 208207 | 125793 | 0.2610 | 28529 | 1.0693 |
| 1992 | 792557 | 217242 | 102884 | 0.2858 | 28985 | 1.0018 |
| 1993 | 580372 | 183534 | 98509 | 0.2482 | 31778 | 0.9912 |
| 1994 | 869256 | 178629 | 89016 | 0.2280 | 24430 | 0.9984 |
| 1995 | 631358 | 159081 | 71728 | 0.2660 | 29575 | 1.0001 |
| 1996 | 835046 | 193038 | 115212 | 0.1703 | 26105 | 1.0477 |
| 1997 | 1491824 | 216559 | 75352 | 0.5070 | 35233 | 1.0079 |
| 1998 | 481352 | 183267 | 98427 | 0.4900 | 33353 | 0.9992 |
| 1999 | 305133 | 141290 | 81928 | 0.3025 | 29736 | 1.0015 |
| 2000 | 1636643 | 199079 | 69449 | 0.2368 | 18322 | 0.9997 |
| 2001 | 1090588 | 221386 | 113982 | 0.2410 | 24556 | 1.0049 |
| 2002 | 1143083 | 255345 | 134943 | 0.2668 | 32914 | 1.0021 |
| 2003 | 434105 | 214036 | 133947 | 0.2315 | 28081 | 1.0074 |
| 2004 | 251977 | 169809 | 119690 | 0.1948 | 25021 | 1.0172 |
| 2005 | 299915 | 140337 | 98238 | 0.1202 | 14129 | 1.0021 |
| 2006 | 554435 | 163066 | 93270 | 0.2272 | 27346 | 0.9997 |
| 2007 | 323159 | 145629 | 91848 | 0.2875 | 29616 | 1.0004 |
| 2008 | 145843 | 120613 | 91884 | 0.1555 | 16054 | 1.0022 |

TABLE 5.6.2.13 HERRING in VIa (N). ESTIMATED FISHING MORTALITY

```
Units : f
    year
age 1957 1958 1959 1960 1961 1962 1962 1963 1964 1965 1966 1967 1968
    1 0.010 0.012 0.040 0.009 0.016 0.038 0.009 0.044 0.062 0.364 0.139 0.088
    2 0.099 0.095 0.107 0.171 0.257 0.261 0.116 0.131 0.069 0.238 0.123 0.146
    3 0.321 0.302 0.157 0.138 0.156 0.206 0.251 0.153 0.147 0.168 0.165 0.139
    4 0.216 0.351 0.398 0.137 0.099 0.243 0.153 0.181 0.178 0.196 0.268 0.164
    5 0.299 0.297 0.353 0.258 0.175 0.168 0.201 0.119 0.203 0.161 0.169 0.137
    6 0.297 0.376 0.309 0.246 0.086 0.205 0.127 0.159 0.104 0.243 0.152 0.130
    7 0.186 0.407 0.372 0.274 0.153 0.215 0.117 0.113 0.199 0.141 0.204 0.100
    8 0.228 0.288 0.269 0.206 0.177 0.235 0.163 0.150 0.144 0.210 0.182 0.146
    9 0.228 0.288 0.269 0.206 0.177 0.235 0.163 0.150 0.144 0.210 0.182 0.146
        year
age 1969 1970 1971 1972 1973 1974 1975 1976 1977 1978 1979 1980
    1 0.020 0.115 0.035 0.367 0.078 0.335 0.138 0.195 0.092 0.040 0.000 0.005
    2 0.081 0. 112 0.416 0.236 0. 502 0.495 0.738 0.774 0.353 0.294 0.001 0.001
    3 0.168 0.349 0.983 0.428 0.587 0.772 0.884 1.219 0.607 0.269 0.001 0.000
    4 0.214 0.401 0.864 0.299 0.629 0.911 0.855 1.086 0.945 0.535 0.000 0.001
    5 0.304 0.375 0.686 0.410 0.594 0.931 0.905 0.892 0.959 0.780 0.000 0.000
    6 0.280 0.307 0.621 0.322 0.612 1.214 0.993 1.074 1.463 1.123 0.002 0.000
    7 0.326 0.457 0.562 0.487 0.354 0.891 1.044 1.107 0.883 1.003 0.002 0.002
    8 0.215 0.315 0.689 0.359 0.573 0.857 0.921 1.042 0.836 0.635 0.001 0.001
    9 0.215 0.315 0.689 0.359 0.573 0.857 0.921 1.042 0.836 0.635 0.001 0.001
    year
age 1981 1982 1983 1984 1985
    1 0.036 0.028 0.044 0.003 0.055 0.062 0.015 0.003 0.012 0.053 0.115 0.011
    2 0.323 0.660 0.391 0.232 0.212 0.551 0.284 0.199 0.138 0.166 0.193 0. 256
    3 0.430 0.604 0.595 0.502 0.308 0.476 0.328 0.311 0.302 0.236 0.156 0.351
    4 0.398 0.806 0.501 0.445 0.306 0.410 0.308 0.333 0.215 0.327 0.242 0.204
    5 0.307 0.720 0.973 0.457 0.314 0.566 0.269 0.282 0.326 0.462 0.243 0.267
    6 0.314 0.570 0.786 0.669 0.335 0.657 0.472 0.217 0.149 0.375 0.403 0.321
    7 0.318 0.465 0.753 0.598 1.014 0.187 0.541 0.298 0.152 0.263 0.338 0.298
    8 0.365 0.678 0.650 0.469 0.391 0.515 0.371 0.276 0.213 0.299 0.264 0. 292
    9 0.365 0.678 0.650 0.469 0.391 0.515 0.371 0.276 0.213 0.299 0.264 0. 292
        year
age 1993 1994 1995 1996 1997 1998 1999 2000 2001 
    1 0.049 0.003 0.001 0.004 0.001 0.030 0.040 0.003 0.001 0.001 0.001 0.000
    2 0.475 0.232 0.351 0.208 0.236 0.170 0.269 0.216 0.121 0.134 0.117 0.098
    3 0.338 0.402 0.328 0.228 0.321 0.238 0.357 0.217 0.236 0.262 0.227 0.191
    4 0.244 0.163 0.365 0.148 0.365 0.515 0.258 0.245 0.222 0.245 0.213 0.179
    5 0.151 0.231 0.193 0.189 0.579 0.434 0.374 0.174 0.271 0.300 0.260 0.219
    6 0.260 0.116 0.178 0.116 0.763 0.773 0.221 0.311 0.235 0.260 0.226 0.190
    7 0.270 0.266 0.160 0.274 0.817 0.977 0.346 0.119 0.283 0.314 0.272 0.229
    8 0.335 0.245 0.294 0.204 0.489 0.483 0.312 0.229 0.222 0.245 0.213 0.179
    9 0.335 0.245 0.294 0.204 0.489 0.483 0.312 0.229 0.222 0.245 0.213 0.179
        year
age 2005 2006 2007 2008
    1 0.000 0.001 0.001 0.000
    2 0.061 0.115 0.145 0.078
    3 0.118 0.223 0.282 0.152
    4 0.111 0.209 0.265 0.143
    5 0.135 0.255 0.323 0.175
    6 0.117 0.222 0.280 0.152
    70.141 0.267 0.338 0.183
    0 0.111 0.209 0.265 0.143
    0 0.111 0.209 0.265 0.143
```

table 5.6.2.14 HERRING in VIa (N). ESTIMATED POPULATION ABUNDANCE


TABLE 5.6.2.15 HERRING in VIa (N). SURVIVORS AFTER TERMINAL YEAR

```
Units : NA
    year
age 2009
    1 NA
    2 53634
    3 81381
    491844
    5 35293
    622336
    729164
    854820
    992560
```

TABLE 5.6.2.16 HERRING in VIa (N). FITTED SELECTION PATTERN

```
Units : NA
    year
age 2001 2002 2003 2004 2005 2006 2007 2008
    1 0.002 0.002 0.002 0.002 0.002 0.002 0.002 0.002
    2 0.548 0.548 0.548 0.548 0.548 0.548 0.548 0.548
    3 1.066 1.066 1.066 1.066 1.066 1.066 1.066 1.066
    41.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    5 1.221 1.221 1.221 1.221 1.221 1.221 1.221 1.221
    6 1.059 1.059 1.059 1.059 1.059 1.059 1.059 1.059
    7 1.277 1.277 1.277 1.277 1.277 1.277 1.277 1.277
    8 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    9 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
```

table 5.6.2.17 herring in via ( N ). PREDICTED CATCH IN Numbers


TABLE 5.6.2.18 HERRING in VIa (N). CATCH RESIDUALS
Units : Thousands NA
year

|  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| 1 | -0.942 | 0.847 | -0.917 | - Inf | 1.288 | -0.284 | 0.008 | - Inf |
| 2 | 0.309 | -0.128 | -0.186 | -0.636 | 0.714 | -0.436 | 0.364 | 0.019 |
| 3 | 0.192 | 0.131 | -0.024 | -0.678 | 0.764 | -0.246 | 0.086 | -0.252 |
| 4 | -0.267 | 0.043 | 0.169 | 0.049 | 0.097 | -0.028 | -0.359 | -0.051 |
| 5 | -0.250 | 0.457 | 0.147 | 0.298 | -0.410 | 0.230 | -0.056 | -0.332 |
| 6 | -0.279 | 0.291 | -0.324 | 0.145 | -0.351 | 0.370 | 0.209 | 0.124 |
| 7 | 0.421 | -0.222 | -0.172 | 0.555 | -0.643 | -0.124 | 0.068 | 0.248 |
| 8 | 0.044 | -0.524 | 0.546 | 0.162 | -0.301 | 0.336 | -0.182 | 0.303 |
| 9 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

TABLE 5.6.2.19 HERRING in VIa (N). PREDICTED INDEX VALUES


TABLE 5.6.2.20 HERRING in VIa (N). INDEX RESIDUALS

WoS Summer Acoustic Survey

| Units : NA |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year |  |  |  |  |  |  |  |  |  |  |  |
| age | 1987 | 198819 | 1989 | 91990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| 1 | -1.008 | NA | NA | A NA | 1.059 | -1.247 | -4.207 | 0.551 | 0.800 | -1.892 | 0.482 |
| 2 | -0.102 | NA | NA | A NA | -0.118 | 0.644 | 0.324 | 0.217 | 0.526 | 0.133 | -0.021 |
| 3 | -0.129 | NA | NA | A NA | -0.819 | -0.418 | 1.001 | 0.303 | 0.119 | 0.263 | -0.543 |
| 4 | -0.368 | NA | NA | A NA | -0.446 | -0.802 | 1.079 | 0.449 | 0.460 | 0.061 | -0.929 |
| 5 | -0.102 | NA | NA | A NA | -0.350 | 0.123 | 0.311 | 0.753 | 0.198 | -0.629 | -0.968 |
| 6 | 0.368 | NA | NA | A NA | 0.187 | -0.656 | 1.239 | -0.146 | 0.579 | -0.459 | -0.665 |
| 7 | -0.157 | NA | NA | A NA | 0.039 | 0.153 | -0.062 | 0.880 | -0.334 | 0.059 | -1.143 |
| 8 | -0.838 | NA | NA | A NA | -0.129 | -0.018 | 1.017 | -0.113 | 0.803 | -0.738 | -0.347 |
| 9 | -2.286 | NA | NA | A NA | -0.149 | -0.169 | 1.258 | -0.232 | 0.419 | -1.675 | -1.176 |
| year |  |  |  |  |  |  |  |  |  |  |  |
| age | 1998 | 1999 |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 1 | 2.062 | 1.696 |  | 0.180 | -0.133 | 0.125 | 1.126 | 1.920 | -0.673 | -0.482 | NA |
| 2 | -0.426 | -0.114 |  | 0.303 | -0.253 | -0.733 | 0.079 | -0.295 | 0.109 | 1.197 | -1.292 |
| 3 | 0.008 | 0.158 |  | 0.074 | -0.089 | 0.001 | -0.034 | -0.322 | -0.607 | 0.478 | 0.114 |
| 4 | 0.457 | -0.102 |  | 0.095 | -0.448 | 0.141 | 0.335 | -0.504 | -0.686 | -0.153 | 0.148 |
| 5 | -0.264 | 0.664 |  | 0.210 | -0.174 | -0.057 | 0.491 | -0.191 | -0.746 | 0.016 | -0.360 |
| 6 | 0.017 | -0.083 |  | 0.903 | -0.553 | 0.179 | 0.096 | -0.349 | -1.239 | 0.081 | -0.115 |
| 7 | -0.221 | 0.777 |  | 0.578 | 0.453 | -0.004 | 0.376 | -0.281 | -1.540 | -0.511 | -0.037 |
| 8 | -0.485 | 0.562 |  | 1.337 | -0.400 | 0.479 | 0.152 | -0.657 | -0.393 | -0.633 | -0.389 |
| 9 | 0.309 | 0.935 |  | 2.097 | 0.653 | 0.943 | 1.045 | -0.163 | -0.666 | -0.497 | -0.690 |
| year |  |  |  |  |  |  |  |  |  |  |  |
| age 2008 |  |  |  |  |  |  |  |  |  |  |  |
| 10.000 |  |  |  |  |  |  |  |  |  |  |  |
| $2-0.176$ |  |  |  |  |  |  |  |  |  |  |  |
| 30.590 |  |  |  |  |  |  |  |  |  |  |  |
| 41.216 |  |  |  |  |  |  |  |  |  |  |  |
| 51.075 |  |  |  |  |  |  |  |  |  |  |  |
| 60.615 |  |  |  |  |  |  |  |  |  |  |  |
| $7 \quad 0.976$ |  |  |  |  |  |  |  |  |  |  |  |
| 80.789 |  |  |  |  |  |  |  |  |  |  |  |
| 9 | 0.044 |  |  |  |  |  |  |  |  |  |  |

TABLE 5.6.2.21 HERRING in VIa (N). FIT PARAMETERS
F, 2001
F, 2002
F, 2003
F, 2004
F, 2005
F, 2006
F, 2007
F, 2008
Selectivity at age 1
Selectivity at age 2
Selectivity at age 3
Selectivity at age 5
Selectivity at age
Selectivity at age 7
Terminal year pop, age 1
Terminal year pop, age 2
Terminal year pop, age 3
Terminal year pop, age 4
Terminal year pop, age 5
Terminal year pop, age 6
Terminal year pop, age 7
Terminal year pop, age 8
Last true age pop, 2001
Last true age pop, 2002
Last true age pop, 2003
Last true age pop, 2004
Last true age pop, 2005
Last true age pop, 2006
Last true age pop, 2007
Index 1, age 1 numbers, $Q$
Index 1, age 2 numbers,
Index $1, ~ a g e ~$
Index $1, ~ a g e ~$

| Value | Std.dev Lower.95.pct.CL Upper.95.pct.CL |  |  |
| ---: | ---: | ---: | ---: |
| 0.22 | 0.15 | 0.16 | 0.30 |
| 0.25 | 0.15 | 0.18 | 0.33 |
| 0.21 | 0.15 | 0.16 | 0.29 |
| 0.18 | 0.16 | 0.13 | 0.25 |
| 0.11 | 0.16 | 0.08 | 0.15 |
| 0.21 | 0.17 | 0.15 | 0.29 |
| 0.26 | 0.20 | 0.18 | 0.39 |
| 0.14 | 0.24 | 0.09 | 0.23 |
| 0.00 | 0.37 | 0.00 | 0.00 |
| 0.55 | 0.15 | 0.41 | 0.73 |
| 1.07 | 0.13 | 0.82 | 1.38 |
| 1.22 | 0.12 | 0.96 | 1.55 |
| 1.06 | 0.12 | 0.84 | 1.33 |
| 1.28 | 0.12 | 1.02 | 1.61 |
| 145841.95 | 2.60 | 885.52 | 24019607.83 |
| 118807.23 | 0.34 | 60791.89 | 232188.20 |
| 130643.13 | 0.27 | 76944.66 | 221816.94 |
| 45001.03 | 0.24 | 27893.19 | 72601.68 |
| 29393.52 | 0.23 | 18868.87 | 45788.59 |
| 37502.81 | 0.22 | 24222.56 | 58064.08 |
| 72729.21 | 0.22 | 47596.02 | 111134.05 |
| 47386.74 | 0.22 | 30675.35 | 73202.19 |
| 23261.08 | 0.28 | 13359.14 | 40502.44 |
| 12599.74 | 0.22 | 8150.90 | 19476.79 |
| 29073.00 | 0.20 | 19578.23 | 43172.40 |
| 46124.03 | 0.19 | 31766.52 | 66970.69 |
| 16181.63 | 0.19 | 11183.39 | 23413.76 |
| 12080.81 | 0.18 | 8539.26 | 17091.18 |
| 73511.93 | 0.19 | 50907.54 | 106153.31 |
| 0.57 | 0.62 | 0.17 | 1.89 |
| 2.87 | 0.19 | 1.97 | 4.17 |
| 4.69 | 0.19 | 3.23 | 6.80 |
| 5.03 | 0.19 | 3.47 | 7.30 |
| 4.58 | 0.19 | 3.16 | 6.65 |
| 4.50 | 0.19 | 3.10 | 6.54 |
| 4.36 | 0.19 | 2.99 | 6.35 |
| 4.00 | 0.19 | 2.73 | 5.87 |
| 4.08 | 0.19 | 2.80 | 5.95 |
|  |  |  |  |

Table 5.7.1.1. Herring in VIa (N). Input data for short-term predictions, numbers at age from the assessment with ages 1 - and 2-ring in 2008 replaced by geometric mean values - natural mortality (M), proportion mature (Mat), proportion of fishing mortality prior to spawning (PF), proportion of natural mortality prior to spawning (PM), mean weights at age in the stock (SWt), selection pattern (Sel), mean weights at age in the catch (CWt). All biological data are taken as mean of the last 3 years. VIa ( N ) herring appears to have considerable annual variability in mean weights and in fraction mature. Last year's values are not applicable. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| 2009 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 1 | 629204 | 1 | 0 | 0.67 | 0.67 | 0.0682 | $3.46 \mathrm{E}-04$ | $6.87 \mathrm{E}-02$ |
| 2 | 231391 | 0.3 | 0.93 | 0.67 | 0.67 | 0.158 | 0.080 | 0.165 |
| 3 | 81381 | 0.2 | 0.99 | 0.67 | 0.67 | 0.180 | 0.155 | 0.187 |
| 4 | 91843 | 0.1 | 1 | 0.67 | 0.67 | 0.195 | 0.147 | 0.207 |
| 5 | 35293 | 0.1 | 1 | 0.67 | 0.67 | 0.200 | 0.176 | 0.216 |
| 6 | 22335 | 0.1 | 1 | 0.67 | 0.67 | 0.204 | 0.156 | 0.232 |
| 7 | 29163 | 0.1 | 1 | 0.67 | 0.67 | 0.211 | 0.187 | 0.238 |
| 8 | 54819 | 0.1 | 1 | 0.67 | 0.67 | 0.217 | 0.147 | 0.269 |
| 9 | 92560 | 0.1 | 1 | 0.67 | 0.67 | 0.228 | 0.147 | 0.289 |


| 2010 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 1 | 629204 | 1 | 0 | 0.67 | 0.67 | 0.0682 | $3.46 \mathrm{E}-04$ | $6.87 \mathrm{E}-02$ |
| 2 | $\cdot$ | 0.3 | 0.93 | 0.67 | 0.67 | 0.158 | 0.080 | 0.165 |
| 3 | $\cdot$ | 0.2 | 0.99 | 0.67 | 0.67 | 0.180 | 0.155 | 0.187 |
| 4 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.195 | 0.147 | 0.207 |
| 5 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.200 | 0.176 | 0.216 |
| 6 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.204 | 0.156 | 0.232 |
| 7 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.211 | 0.187 | 0.238 |
| 8 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.217 | 0.147 | 0.269 |
| 9 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.228 | 0.147 | 0.289 |


| 2011 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 1 | 629204 | 1 | 0 | 0.67 | 0.67 | 0.0682 | $3.46 \mathrm{E}-04$ | $6.87 \mathrm{E}-02$ |
| 2 | $\cdot$ | 0.3 | 0.93 | 0.67 | 0.67 | 0.158 | 0.080 | 0.165 |
| 3 | . | 0.2 | 0.99 | 0.67 | 0.67 | 0.180 | 0.155 | 0.187 |
| 4 | . | 0.1 | 1 | 0.67 | 0.67 | 0.195 | 0.147 | 0.207 |
| 5 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.200 | 0.176 | 0.216 |
| 6 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.204 | 0.156 | 0.232 |
| 7 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.211 | 0.187 | 0.238 |
| 8 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.217 | 0.147 | 0.269 |
| 9 | $\cdot$ | 0.1 | 1 | 0.67 | 0.67 | 0.228 | 0.147 | 0.289 |

Table 5.7.1.2. Herring in VIa (N). Short-term prediction single option table, with TAC constraint. N.B. In this table "age" refers to number of rings (winter rings in the otolith).

| Year: | 2009 | F multiplier: | 1.45 | Fbar: | 0.23 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0003 | 200 | 14 | 629205 | 42912 | 0 | 0 | 0 | 0 |
| 2 | 0.08 | 21971 | 3627 | 231391 | 36606 | 215194 | 34044 | 162841 | 25761 |
| 3 | 0.16 | 14943 | 2798 | 81381 | 14657 | 80432 | 14486 | 60483 | 10893 |
| 4 | 0.15 | 16861 | 3489 | 91844 | 17900 | 91844 | 17900 | 74423 | 14505 |
| 5 | 0.18 | 7600 | 1644 | 35293 | 7075 | 35293 | 7075 | 27805 | 5574 |
| 6 | 0.16 | 4325 | 1003 | 22336 | 4564 | 22336 | 4564 | 17941 | 3666 |
| 7 | 0.19 | 6622 | 1575 | 29164 | 6160 | 29164 | 6160 | 22731 | 4802 |
| 8 | 0.15 | 10064 | 2708 | 54820 | 11907 | 54820 | 11907 | 44422 | 9648 |
| 9 | 0.15 | 16993 | 4904 | 92560 | 21082 | 92560 | 21082 | 75004 | 17083 |
| Total |  | 99580 | 21760 | 1267992 | 162863 | 621641 | 117218 | 485649 | 91933 |
| 7 | 0.15 | 13029 | 31336 | 17252 | 1349209 | 176979 | 701961 | 131173 | 568631 |
| Total |  |  |  |  |  |  |  |  |  |

Table 5.7.1.3. Herring in VIa (N). Short-term prediction multiple option table, with TAC constraint.

| 2008 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BIomass | SSB | FMULT | FBAR | Landings |  |  |  |
| 162863 | 91933 | 1.4512 | 0.2306 | 21760 |  |  |  |
| 2009 |  |  |  |  | $2010$ |  |  |
| BIOMASS | SSB | FMULT | FBAR | LANDINGS | Biomass | SSB | \% <br> CHANGE |
| 168163 | $108800$ | 0 | 0 | 0 | 191031 | 129944 | -100\% |
| . | $107807$ | $0.1$ | $0.0159$ | $1718$ | $189533$ | 127373 | -92\% |
| . | $106825$ | $0.2$ | $0.0318$ | $3413$ | 188057 | 124861 | $-84 \%$ |
| . | 105852 | 0.3 | 0.0477 | 5083 | 186601 | 122407 | -77\% |
| . | $104888$ | 0.4 | $0.0636$ | $6730$ | $185167$ | 120010 | -69\% |
| . | $103933$ | $0.5$ | $0.0794$ | $8354$ | 183753 | 117668 | -62\% |
| . | $102988$ | $0.6$ | $0.0953$ | $9955$ | 182359 | 115380 | $-54 \%$ |
| . | 102052 | $0.7$ | 0.1112 | 11534 | 180985 | 113144 | -47\% |
| . | $101125$ | $0.8$ | $0.1271$ | $13090$ | 179630 | 110960 | -40\% |
| . | $100207$ | $0.9$ | $0.143$ | $14625$ | 178295 | 108826 | -33\% |
| . | $99298$ | 1 | $0.1589$ | $16139$ | 176979 | 106740 | $-26 \%$ |
| . | $98398$ | $1.1$ | $0.1748$ | $17631$ | 175682 | 104703 | -19\% |
| . | 97507 | $1.2$ | $0.1907$ | 19103 | 174403 | 102712 | -12\% |
| . | 96624 | $1.3$ | $0.2066$ | $20554$ | 173143 | 100766 | $-6 \%$ |
| . | $95749$ | $1.4$ | 0.2224 | 21985 | 171900 | 98864 | 1\% |
| . | $94883$ | $1.5$ | 0.2383 | 23396 | 170675 | 97006 | 8\% |
| . | 94026 | $1.6$ | 0.2542 | 24788 | 169467 | 95189 | 14\% |
| . | $93176$ | $1.7$ | $0.2701$ | 26160 | 168277 | 93414 | $20 \%$ |
| . | 92335 | $1.8$ | 0.286 | 27513 | 167103 | 91679 | 26\% |
| . | 91502 | $1.9$ | $0.3019$ | 28848 | 165946 | 89983 | 33\% |
| . | 90677 | 2 | 0.3178 | 30164 | 164806 | 88325 | $39 \%$ |
| Values for catch option table |  |  |  |  |  |  |  |
| 168163 | 108800 | 0 | 0 | 0 | 191031 | 129944 | -100\% |
| . | 98398 | $1.1$ | 0.17 | $17631$ | $175682$ | $104703$ | -20\% |
| . | $95749$ | $1.4$ | $0.22$ | $21760$ | $171900$ | $98864$ | $0 \%$ |
|  | 94252 | $1.6$ | $0.25$ | $24420$ | $169786$ | 95669 | $12 \%$ |
|  | 93181 | 1.7 | 0.27 | 26151 | $168284$ | $93425$ | 20\% |
|  | 92520 | $1.8$ | $0.28$ | $27215$ | $167361$ | $92061$ | $25 \%$ |
|  | $91602$ | $1.9$ | $0.30$ | $28688$ | $166084$ | $90186$ | $32 \%$ |
|  | 89006 | 2.2 | 0.35 | 32829 | 162497 | 84967 | 51\% |

Table 5.8.1.1. Herring in VIa (N). Covariance matrix from ICA used to provide uncertainty of initial numbers for medium term simulations.

| age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.117 | 0.008 | 0.009 | 0.008 | 0.007 | 0.007 | 0.007 | 0.007 | 0.000 |
| 2 | 0.008 | 0.117 | 0.047 | 0.042 | 0.037 | 0.037 | 0.034 | 0.032 | 0.007 |
| 3 | 0.009 | 0.047 | 0.073 | 0.039 | 0.036 | 0.035 | 0.032 | 0.031 | 0.032 |
| 4 | 0.008 | 0.042 | 0.039 | 0.060 | 0.036 | 0.035 | 0.033 | 0.033 | 0.031 |
| 5 | 0.007 | 0.037 | 0.036 | 0.036 | 0.051 | 0.035 | 0.034 | 0.034 | 0.033 |
| 6 | 0.007 | 0.037 | 0.035 | 0.035 | 0.035 | 0.050 | 0.036 | 0.036 | 0.034 |
| 7 | 0.007 | 0.034 | 0.032 | 0.033 | 0.034 | 0.036 | 0.047 | 0.037 | 0.036 |
| 8 | 0.007 | 0.032 | 0.031 | 0.033 | 0.034 | 0.036 | 0.037 | 0.049 | 0.037 |
| 9 | 0.000 | 0.007 | 0.032 | 0.031 | 0.033 | 0.034 | 0.036 | 0.037 | 0.049 |

Table 5.8.1.2. Herring in VIa (N). Parameters of $S / R$ relationships.

| Type | equation | a | b | sigma(ln) |
| :--- | :--- | :--- | :--- | :--- |
| B/H 1957-2006 | AS/(B+S) | 4475 | 403 | 0.61 |
| B/H 1989-2006 | AS/(B+S) | 3260 | 403 | 0.477 |
| H-S 1989-2006 | A(B-S)/B $\ldots . . \mathrm{S}<\mathrm{B}$ | 637 | 50 | 0.52 |
|  | $\mathrm{~A} \ldots \ldots . . . . . \mathrm{S} \geq \mathrm{B}$ |  |  |  |

Table 5.8.1.3. Herring in VIa (N). Medium term run details.

| Run <br> Number | Period for <br> S/R data | S/R Model | $\%$ y-y TAC <br> constraint <br> Blim<SSB<Btrig | $\%$ random <br> Measurement <br> error | $\%$ <br> implementation <br> bias |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | $89-2006$ | B/H | $20 \%$ | $25 \%$ | $0 \%$ |
| 2 | $89-2006$ | B/H | $25 \%$ | $25 \%$ | $0 \%$ |
| 3 | $89-2006$ | B/H | $20 \%$ | $30 \%$ | $0 \%$ |
| 4 | $89-2006$ | B/H | $25 \%$ | $30 \%$ | $0 \%$ |
| 5 | $89-2006$ | H-S | $20 \%$ | $25 \%$ | $0 \%$ |
| 6 | $89-2006$ | H-S | $25 \%$ | $25 \%$ | $0 \%$ |
| 7 | $89-2006$ | H-S | $20 \%$ | $30 \%$ | $0 \%$ |
| 8 | $89-2006$ | H-S | $25 \%$ | $30 \%$ | $0 \%$ |
| 9 | $89-2006$ | H-S | $25 \%$ | $25 \%$ | $10 \%$ |
| 10 | $89-2006$ | H-S | $25 \%$ | $25 \%$ | $20 \%$ |
| 11 | $89-2006$ | H-S | $25 \%$ | $25 \%$ | $30 \%$ |



Figure 5.1. Location of ICES area VIa (North) and adjacent areas, with place names.


Figure 5.6.1.1. Herring in VIa (N). Mean squared deviance ( y axis) between previous and current assessments for the 8 assessment years in the retrospective and years ( $x$ axis) previous to the terminal year in the assessment. 2008 WG data with flat weighting at age. SSB (dots) and ${ }_{3}{ }_{36}$ (solid line). Across rows selection at oldest age ranging from 0.7 to 1.2, down columns weighting on catch from 2.0 to 0.5 . Minimum is candidate for best settings (see text table section 5.6.1)


Figure 5.6.1.2. Herring in VIa (N). Mean squared deviance (y axis) between previous and current assessments for the 8 assessment years in the retrospective and years ( $x$ axis) previous to the terminal year in the assessment. 2008 WG data with varying weighting at age. SSB (dots) and F $_{3-6}$ (solid line). Across rows selection at oldest age ranging from 0.7 to 1.2 , down columns weighting on catch from 2.0 to 0.5 . Minimum is candidate for best settings (see text table section 5.6.1)


Figure 5.6.1.3. Herring in VIa (N). Mean squared deviance (y axis) between previous and current assessments for the 8 assessment years in the retrospective and years ( x axis) previous to the terminal year in the assessment. 2009 WG data with flat weighting at age. SSB (dots) and F3-6 (solid line). Across rows selection at oldest age ranging from 0.7 to 1.2 , down columns weighting on catch from 2.0 to 0.5 . Minimum is candidate for best settings (see text table section 5.6.1)


Figure 5.6.1.4. Herring in VIa (N). Mean squared deviance (y axis) between previous and current assessments for the 8 assessment years in the retrospective and years ( $x$ axis) previous to the terminal year in the assessment. 2009 WG data with varying weighting at age. SSB (dots) and $\mathrm{F}_{3-6}$ (solid line). Across rows selection at oldest age ranging from 0.7 to 1.2, down columns weighting on catch from 2.0 to 0.5 . Minimum is candidate for best settings (see text table section 5.6.1)
a) SPALY









Figure 5.6.1.5. Herring in VIa (N). Comparison between retrospective performance with a) SPALY settings and b) 2009 optimal choices of model settings and fit diagnostics for catch under both circumstances. The fit is very similar and although the restrospective performance can be improved, the settings are not stable across years so no changes are suggested (see section 5.6.1)


Figure 5.6.2.1. Herring in VIa (N). Illustration of stock trends from the assessment (8 year separable period) 1957-2008. Summary of estimates of landings, spawning stock biomass at spawning time, fishing mortality at $\mathrm{F}_{3-6}$, recruitment at 1-ring, in the final assessment run.

## Fitted catch diagnostics



Figure 5.6.2.2. Herring in VIa (N). Illustration of selection patterns diagnostics, from deterministic calculation (8-year separable period). Top left, a bubble plot of selection pattern residuals. Top right, estimated selection (relative to 4-ringers) +/- standard deviation. Bottom, marginal totals of residuals by year and ring.


Figure 5.6.2.3. Herring in VIa (N). Illustration of residuals from deterministic calculation (8-year separable period). Diagnostics of the fit of the 1-ring index against the acoustic surveys. Top left, fitted populations (line), and predictions of abundance in each year made from the index observations and estimated catchability (triangles +/- standard deviation), plotted by year. Top right, scatter plot and fitted relationship of abundance from fitted populations of 1 -ringers in acoustic surveys. Bottom, residuals, as $\ln$ (observed index) - $\ln$ (expected index) plotted against expected values and against time. N.B. 1-ringers are down-weighted in the catch and survey in the assessment.


Figure 5.6.2.4. Herring in VIa (N). Illustration of residuals from deterministic calculation (8-year separable period). Diagnostics of the fit of the 2-ring index against the acoustic surveys. Top left, fitted populations (line), and predictions of abundance in each year made from the index observations and estimated catchability (triangles +/- standard deviation), plotted by year. Top right, scatter plot and fitted relationship of abundance from fitted populations of 2-ringers in acoustic surveys. Bottom, residuals, as $\ln (o b s e r v e d ~ i n d e x) ~-~ l n(e x p e c t e d ~ i n d e x) ~ p l o t t e d ~ a g a i n s t ~ e x p e c t e d ~$ values and against time.



Figure 5.6.2.6. Herring in VIa (N). Illustration of residuals from deterministic calculation (8-year separable period). Diagnostics of the fit of the 4 -ring index against the acoustic surveys. Top left, fitted populations (line), and predictions of abundance in each year made from the index observations and estimated catchability (triangles +/- standard deviation), plotted by year. Top right, scatter plot and fitted relationship of abundance from fitted populations of 4 -ringers in acoustic surveys. Bottom, residuals, as $\ln (o b s e r v e d ~ i n d e x) ~-~ l n(e x p e c t e d ~ i n d e x) ~ p l o t t e d ~ a g a i n s t ~ e x p e c t e d ~$ values and against time.


Figure 5.6.2.7. Herring in VIa (N). Illustration of residuals from deterministic calculation (8-year separable period). Diagnostics of the fit of the 5 -ring index against the acoustic surveys. Top left, fitted populations (line), and predictions of abundance in each year made from the index observations and estimated catchability (triangles +/- standard deviation), plotted by year. Top right, scatter plot and fitted relationship of abundance from fitted populations of 5 ringers in acoustic surveys. Bottom, residuals, as $\ln$ (observed index) - $\ln$ (expected index) plotted against expected values and against time.


Figure 5.6.2.8. Herring in VIa (N). Illustration of residuals from deterministic calculation (8-year separable period). Diagnostics of the fit of the 6-ring index against the acoustic surveys. Top left, fitted populations (line), and predictions of abundance in each year made from the index observations and estimated catchability (triangles +/- standard deviation), plotted by year. Top right, scatter plot and fitted relationship of abundance from fitted populations of 6 ringers in acoustic surveys. Bottom, residuals, as $\ln ($ observed index) - $\ln$ (expected index) plotted against expected values and against time.


Figure 5.6.2.9. Herring in VIa (N). Illustration of residuals from deterministic calculation (8-year separable period). Diagnostics of the fit of the 7-ring index against the acoustic surveys. Top left, fitted populations (line), and predictions of abundance in each year made from the index observations and estimated catchability (triangles +/- standard deviation), plotted by year. Top right, scatter plot and fitted relationship of abundance from fitted populations of 7 ringers in acoustic surveys. Bottom, residuals, as $\ln (o b s e r v e d ~ i n d e x) ~-~ l n(e x p e c t e d ~ i n d e x) ~ p l o t t e d ~ a g a i n s t ~ e x p e c t e d ~$ values and against time.


Figure 5.6.2.10. Herring in VIa (N). Illustration of residuals from deterministic calculation (8-year separable period). Diagnostics of the fit of the 8 -ring index against the acoustic surveys. Top left, fitted populations (line), and predictions of abundance in each year made from the index observations and estimated catchability (triangles +/- standard deviation), plotted by year. Top right, scatter plot and fitted relationship of abundance from fitted populations of 8 ringers in acoustic surveys. Bottom, residuals, as $\ln ($ observed index) - $\ln$ (expected index) plotted against expected values and against time.


West of Scotland Herring Weighted Residuals Bubble Plot


Figure 5.6.2.12. Herring in VIa (N). Comparison of residuals in the catch (top) and survey (bottom) Note the year effects in the survey, particularly in 2005 and 2008. The assessment effectively smoothes an otherwise noisy survey.

West of Scotland Herring Retrospective Summary Plot


Figure 5.6.2.13. Herring in VIa (N). Analytical retrospective patterns (2008 to 2001) of SSB, mean $F_{3-6}$ and recruitment from the final assessment.



MFYPR version 2 a
Run: TAC
Time and date: 13:17 17/03/2009

| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(3-6) | 1.0000 | 0.1589 |
| FMax | 200.9425 | 31.9275 |
| F0.1 | 1.0811 | 0.1718 |
| F35\%SPR | 1.1285 | 0.1793 |

Weights in kilograms
Figure 5.7.2.1. Herring in VIa (N). Yield-per-recruit and short-term forecast.

MFDP version 1a
Run: TAC
West of Scotland Herring
Time and date: 11:08 19/03/2009
Fbar age range: 3-6
Input units are thousands and kg - output in tonnes

Stock Recruit


AR(1) Residuals


Res iduals by Estimated Recruits


Residuals by year


Residuals by SSB


Normal Q-Q Plot


Figure 5.8.1.1. Herring in VIa (N). Fitted Stock / Recruit relationship for full time series, showing slope used for truncated period 1989 onwards


Figure 5.8.1.2. Herring in VIa (N). Fitted Stock / Recruit model to truncated data 1989 -2006 showing no point of inflection of recruitment with SSB. The model used to conform to this fit is a hockey-stick with a point of inflection at lowest observed biomass of 50000 t .


Figure 5.8.1.3. Herring in VIa (N). Comparison of observed (1989-2006) recruitment and medium term simulations with Hockey-Stick Stock / Recruit relationship.




Figure 5.8.2.1. Herring in VIa (N). Results of medium term simulations of EC harvest rule (Section 5.1.3), model parameters in Section 5.8. Results show a) SSB (Risk SSB<Blim), b) Catch, c) F and d) R from 2009 to 2019. Two models are similar with an initial risk (a) reducing with time, more slowly with Beverton-Holt model than Hockey-Stick model. Risk reduces to low level in both cases.



Figure 5.8.2.2. Herring in VIa (N). Risk of SSB< Blim under different model conditions. Results of medium term simulations of EC harvest rule (see Section 5.1.3.), for model parameters given in section 5.8. a) Sensitivity to $20 / 25 \%$ year-on-year constraint on TAC change (dotted / solid lines). Sensitivity to choice of S/R model; Hockey-Stick (runs 9, 10, 11, 12) Beverton-Holt (runs 5, 6, 7, 8). Sensitivity to choice of measurement error (CV of $25 \%$ runs $9,10,5,6$ or $\mathbf{3 0 \%}$ runs $11,12,7,8$ ) b) Sensitivity to over catch $0,10,20,30 \%$ bias for the case of the HS model runs $6,13,14,15$. In all cases risks fall below $5 \%$ but at $30 \%$ bias (overcatch) risks are no longer negligible.

## West of Scotland Herring Otolith Plot



Fbar (3-6)

Figure 5.10.1. Herring in VIa (N). Model uncertainty; distribution and quantiles of estimated SSB and $F$ in the terminal year of the assessment. Estimates of precision are based on a parametric bootstrap from the FLICA estimated variance/covariance estimates from the model.

## 6 Herring in Divisions Vla (South) and VIIb,c

This management unit has existed since 1982 when it was separated from VIa. Until that time, VIIb,c was a separate management unit. The stock area comprises autumn and winter, and spring spawning components. This stock is classified as "SALY" in 2009.

### 6.1 The Fishery

### 6.1.1 Advice and management applicable to 2008-2009

The TAC for this area in 2008 was 11642 t with a decrease of $20 \%$ to 9314 t in 2009 . For 2009, ICES advised that the updated exploratory assessment available for this stock did not change the perception of the stock and did not give reason to change the advice from 2007. The advice for the fishery in 2009 is therefore the same as the advice given in recent years. A rebuilding plan should be put in place that will reduce catches. The rebuilding plan should be evaluated with respect to the precautionary approach.

### 6.1.2 Catches in 2008

The working group estimates of landings recorded by each country from this fishery from 1988-2008 are given in Table 6.1.2.1 Irish catch estimates for this WG have been based on the preliminary official reported data from the EU Logbook Scheme. The total official catch recorded from logbooks for 2008 was over 10237 t , compared with 12675 t in 2007. The total working group estimates of catches in these areas from 1970 -2008 are shown in Figure 6.1.2.1. The working group estimates of catch have declined from about 18000 t in 2007 to 13000 t in 2008.

There were no estimates of discards reported for 2008 and anecdotal reports from the industry are that there was some discarding in 2008. Some slippage took place but it is not possible to quantify exact amounts.

The assessment period runs concurrently with the annual quota. In recent years Ireland is the dominant country participating in this fishery. In 2008 all of the catches were reported from quarters 1 and 4 in VIaS with comparatively small catches reported from VIIb, c. In the first quarter the season around the $7^{\text {th }}$ of January and closed on the $7^{\text {th }}$ February. Fishing reopened in the fourth quarter on the $7^{\text {th }}$ October and closed on the $15^{\text {th }}$ of December when the quota was exhausted. The distribution of the landings from this area are presented in Figure 6.1.3.1. The main fishing took place in the northern part of VIaS, with most of the remainder from the southern part. There was very little fishing in VIIbc. Several small landings were taken from the west of Ireland, and almost no fishing took place in the northern part of VIIbc.

A total of 62 boats, categorised as follows caught herring in 2008:

- 1 freezer trawler
- 22 pelagic segment boats with refrigerated seawater (RSW) storage
- 4 polyvalent segment boats with refrigerated seawater storage
- 35 polyvalent segment vessels with bulk storage.

Polyvalent is a term used to define part of the Irish fleet licensed to catch pelagic and demersal fish.

### 6.1.3 Regulations and their effects

Changes to the management of this stock have influenced the way the fishery is prosecuted in space and time. The RSW vessels do not have access to those spawning grounds within a 12-mile limit. Fish on the spawning grounds are targeted largely by dry hold vessels only.

The quota is allocated to the RSW and polyvalent vessels in different ways. The RSW vessels are given quota on a fixed allocation key. The polyvalent vessels need to "book in" to receive the remaining quota, and must take it in a specified time window. Unused quota is re-assigned to the next time window. This leads to wastage of quota that is not caught and at least partly explains why the Irish quota was not fully taken up in 2008.

### 6.1.4 Changes in fishing technology and fishing pattern

There have been no significant changes in the fishing technology of the fleets in this area in the very recent past. The pattern of this fishery has changed over time. In the early part of the 20th century the main spawning components were the winter spawners off the north coast, and this was where the main fishery took place. In the 1970s and 1980s the west of Ireland autumn-spawning components were dominant and the fishery was mainly distributed along the coasts of VIIbc and VIaS. More recently the northern grounds are more important again.
The pelagic segment vessels are not allowed to fish herring inside 12 nautical miles of the Irish coast. This means that they tend to fish off the north coast, where the waters are less deep and where herring spawning grounds are further offshore. This exclusion is more enforced than previously.

### 6.2 Biological composition of the catch

### 6.2.1 Catch in numbers-at-age

Catch-at-age data for this fishery are available since 1970 and are shown in Table 6.2.1.1 with percentages since 1994 shown in Table 6.2.1.2. In 2008 the fishery has been dominated by 2,3 and 4 ringers, accounting for $15 \%, 24 \%$ and $35 \%$ respectively. One ringers are never well represented in the catch and normally do not show up in the catch until quarter 3 . In any case, the abundance of 1-ringer in the catches has been lower in the past three years than at any time in the series. The 2008 age profile shows a a peak in 4 ringers. This follows on from the strong catch of 3 ringers in 2007. The catch numbers at age have been mean standardised and are presented in Figure 6.2.1.1. The low numbers of 1 ringers and the truncation of older ages can be clearly seen.

Four winter ring fish dominate the catch in quarter 1 while in quarter $4,2,3$ and 4 ringers are found in similar quantities. Sampling data indicates that herring are fully recruited to the fishery at 3 ringers and there is little evidence for 1 ringer fish being an important component of landings in fisheries in this area.

### 6.2.2 Quality of the catch and biological data

The management of the Irish fishery in recent years has tightened considerably and the accuracy of reported catches is also believed to have improved. The numbers of samples and the associated biological data are shown in Table 6.2.2.1. As Ireland is the main participant in this fishery all of the sampling is carried out by Ireland. The length distributions of the catches taken per quarter by the Irish fleet are shown in Table 6.2.2.2. Only one sample was collected from VIIb, and overall landings in this area are very small. Sampling in this fishery relies heavily on the vessels that concentrate their effort on the inshore grounds.

### 6.3 Fishery Independent Information

### 6.3.1 Acoustic Surveys

In 2008, the Irish survey of VIaS, VIIb, c was conducted in July with effort concentrating on summer feeding aggregations. The July 2008 survey track and SA values attributed to herring are shown in Figures 6.3.2.1 and 6.3.2.2 respectively. A primary survey was carried out on the Celtic Explorer and a scoping survey was conducted using a commercial vessel. The purpose of the commercial survey was to determine the extent and distribution of summer feeding aggregations of herring within the survey confines.

The main survey focused on the northwest and west coast of Ireland (ICES Divisions VIaS and VIIb, c). The survey track commenced off the west coast of Ireland at the south-eastern extension and worked in continuity from south to north. Existing survey methodology was followed with acoustic surveying undertaken between 04:00 and 23:00 (daylight hours). The commercial survey focused effort in the ICES areas VIIb and VIaS offshore from known autumn and winter spawning grounds where the fishery is focused in quarters 4 and 1 . The vessels also covered the grounds extending to the shelf break and northwards to the $56^{\circ} \mathrm{N}$

A systematic parallel transect design was adopted with a randomised start point. As this was the first of a new survey time series it was deemed important to cover the grounds as intensively as possible to highlight any potential areas of distribution or dense aggregations.

The results of this acoustic survey are not directly comparable with the winter surveys conducted from 2004-2007 (Table 6.3.1.1). It is comparable in time and area with those conducted from 1994-1996 (Table 6.3.1.2). The SSB estimate ( $43,000 \mathrm{t}$ ) was lower than surveys in 1994 and 1995, though in the same range of SSB estimates as the winter surveys conducted in recent years. It remains unclear if the VIaS and VIIbc stock is contained within the area of this survey as herring abundance increased moving towards the boundary with VIaN. This survey is now conducted as part of the PGIPS survey programme (ICES, 2009, LRC:02).

### 6.4 Mean weights-at-age and maturity-at-age

### 6.4.1 Mean Weights at Age

The mean weights ( kg ) at age in the catches in 2008 are based on Irish catches and are very similar to 2007 for ringers 2-6 (Figure 6.4.1.1). These mean weights display quite a stable pattern over the time series. Though there appears to be a slight increase in mean weights of 1 ringers in the past five years. Fluctuations can also be seen in the
oldest two ages with a decrease in 8 ringers and a slight increase in 9 ringers in 2008. Generally the oldest and youngest ages are poorly represented in the catch data.
The mean weights in the stock at spawning time have been calculated from Irish samples taken during the main spawning period that extends from October to February (Figure 6.4.1.2). There appears to be a slight decrease in 8 ringers, an increase in $9+$ and a more stable pattern in the younger ages.

### 6.4.2 Maturity Ogive

One ringers are considered to be immature and they do not contribute to the SSB. This corresponds with the constant maturity ogive that is assumed for this stock and used in the assessment.

### 6.5 Recruitment

There is little information on recruitment in the catch at age data and there are as yet no recruitment indices from the surveys. Numbers of 1-ringers in the catches vary widely but have been consistently low in the most recent years.

### 6.6 Stock Assessment

### 6.6.1 Data Exploration

A detailed analysis of basic data, including age composition of catches, log catch ratios and cohort catch curves was conducted in recent years and is presented in the Stock Annex (annex 7). There has been a truncation in older age groups in recent years, and in most recent years, a paucity of recruits also. Log catch ratios show an upward trend in raw mortality on fully recruited year classes, since the mid 1990s. Catch curves show low mortality on the very large 1981, 1985 and 1988 year classes. These represent three of the biggest year classes recruited to this fishery. Low mortality was evident in the 1970s and increased mortality can be seen from 1990 on.

### 6.6.2 Assessment

Following the procedure of recent years, a separable VPA was used to screen over four terminal fishing mortalities, $0.2,0.4,0.5$ and 0.6 . This was achieved using the Lowestoft VPA software (Darby and Flatman, 1994). Reference age for calculation of fishing mortality was 3-6 and terminal selection was fixed at 1, relative to 3 winter rings. This assessment is still exploratory, and no assessment has been accepted in recent years.

Four assessments using the separable VPA are presented, based on the four choices of terminal F. Recruitment, SSB and mean F from each run are plotted in Figure 6.6.2.1. This figure is more informative for the converged part of the VPA, but in most recent years has little information on the current stock dynamics. Outputs from separable VPAs with terminal Fs of $0.2,0.4,0.5$ and 0.6 are presented in Tables 6.6.2.1, 6.6.2.2, 6.6.2.3 and 6.6.2.4 respectively. Residual plots for the four trial assessments are presented in Figure 6.6.2.2. Large residuals can be seen in 1 ringers. A comparison with the previous year's separable VPA runs is shown in Figure 6.6.2.3.
Fishing mortality was highest in series in 1998. Subsequent Fs have been lower but still above the long term average in each case. There was a sharp rise in F in 2006, associated with an increased catch in that year.

Recruitment appears to have shown a declining trend over the last few years with all terminal F values used. A slightly higher level of recruitment is estimated with terminal $\mathrm{F}=0.2$. Each scenario shows recruitment to be at a similar level in the final year and this is calculated using the geometric mean of the recruitment index over the entire time series.

SSB may be declining slightly, assuming terminal F of $0.4,0.5$ or 0.6 and possibly more stable at F values of 0.2 . All F values show that SSB at lowest levels in the series and is considerably lower than the current levels of $\mathrm{B}_{\mathrm{pa}}$ and $\mathrm{Blim}_{\mathrm{lim}}$. There is no evidence in the observed catch numbers at age to suggest that there are strong year classes recruiting to this fishery.

These explorations are only useful as indicators of historic trends. These results are consistent with the preliminary data screening that shows no stronger year classes in the fishery in recent years.

A retrospective assessment was conducted for each of the F scenarios. Using a terminal F $=0.2$ (Figure 6.6.2.4) overestimates SSB and underestimates F. Using a terminal $\mathrm{F}=0.4$ or 0.5 (Figure 6.6.2.5 and 6.6.2.6) displays a much more stable estimation of SSB and the underestimation of $F$ is not as pronounced. The retrospective assessment using $\mathrm{F}=0.6$ (Figure 6.6.2.7) shows SSB to be quite stable, with some tendency to underestimate in most years, and a tendency for mean F to be overestimated.

The results of the retrospective analysis suggest that using a terminal F of 0.4 or 0.5 produces more stable estimates of SSB and F than smaller or larger values. This suggests that recent F has been in the range of 0.4 to 0.5 , which is above $\mathrm{F}_{0.1}$ estimated most recently in 2006.

### 6.6.3 State of the Stock

The results of the exploratory assessment suggest that the decline in SSB may be continuing but the current level of SSB is uncertain but is likely to be below $\mathrm{B}_{\mathrm{pa}}$ and $\mathrm{Blim}_{\mathrm{lim}}$. There is no evidence that large year classes have recruited to the stock in recent years and F appears to have been reduced due to the decrease in catch. The perception of stock trends is consistent, even though the most recent estimates of SSB and F are uncertain.

### 6.7 Short term projections

In the absence of an agreed assessment, it was not considered informative to carry out any predictions.

### 6.8 Medium term projections

Yield per recruit analyses were performed in 2006, and it is not considered necessary to update them. The results from this yield per recruit show $\mathrm{F}_{0.1}=0.17$.

### 6.9 Precautionary and yield based reference points

As this assessment is still uncertain there was no revision of the precautionary reference points. The precautionary reference points for this stock were discussed in the 1999 Working Group Report (ICES 1999 ACFM:12). The present analysis, although uncertain, presents a similar picture of the stock as that shown in recent years. The SGPRP (ICES 2003/ACFM: 15) has reviewed the methodology for the calculation of biological reference points, and applying a segmented regression to the stock and
recruit data from the 2002 HAWG assessment showed that the fit to the stock and recruit data for this stock was not significant. The stock is still likely below Bpa (110 000 t ) but the fishing mortality has been relatively stable, over the past number of years.

### 6.10 Quality of the Assessment

The assessment presented was based on the results from a separable VPA without a tuning index, therefore the estimates of SSB and F for recent years depend on the choice of terminal F. Although landings seem to have been declining in recent years the actual F cannot be determined. Therefore the VPA was run for a range of terminal F values and the current perception of the stock would be highly influenced by that choice. There is no information on recent recruitment levels both because the selectivity of the fishery appears to be low for the juveniles and also the lack of a recruitment index.

The retrospective analysis of the assessment suggests that an F of 0.2 underestimates mean F and SSB. Using the terminal $\mathrm{F}=0.4$ or $\mathrm{F}=0.5$ produces a more stable retrospective pattern. The highest F of 0.6 used shows an overestimation of F. Based on this information we can infer that F may be in the region of $0.4-0.5$.

### 6.11 Management Considerations

The current catch regime which has been in place for a number of years does not appear to have reduced F below Flim. SSB may be stable at an historical low level, and is declining further in all runs. Though little information on recruitment is available, it is unlikely that it is above average and more likely below average. Certainly every effort should be taken to maintain catches below the current level. The catch target of the local management plan is not likely to be achievable at current stock productivity.
Recent mean $F$ may be well above $F_{0.1}$ and this suggests that $F$ and catch need to be reduced. A rebuilding plan is urgently required and should include further substantial reductions in catches.

### 6.12 Environment

### 6.12.1 Ecosystem Considerations

No new information.

### 6.12.2 Changes in the Environment

No new information.

Table 6.1.2.1. VIa(S) and VIIb,c herring. Estimated Herring catches in tonnes, 1988-2008. These data do not in all cases correspond to the official statistics and cannot be used for management purposes.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | - | - | + | - | - | - | - | - | - | - |
| Germany, <br> Fed.Rep. | - | - | - | - | 250 | - | - | 11 | - | - |
| Ireland | 15000 | 18200 | 25000 | 22500 | 26000 | 27600 | 24400 | 25450 | 23800 | 24400 |
| Netherlands | 300 | 2900 | 2533 | 600 | 900 | 2500 | 2500 | 1207 | 1800 | 3400 |
| UK <br> (N.Ireland) | - | - | 80 | - | - | - | - | - | - | - |
| UK <br> (England + <br> Wales) | - | - | - | - | - | - | 50 | 24 | - | - |
| UK Scotland | - | + | - | + | - | 200 | - | - | - | - |
| Total landings | 15300 | 21100 | 27613 | 23100 | 27150 | 30300 | 26950 | 26692 | 25600 | 27800 |
| Unallocated/ area misreported | 13800 | 7100 | 13826 | 11200 | 4600 | 6250 | 6250 | 1100 | 6900 | -700 |
| Discards | - | 1000 | 2530 | 3400 | 100 | 250 | 700 | - | - | 50 |
| WG catch | 29100 | 29200 | 43969 | 37700 | 31850 | 36800 | 33900 | 27792 | 32500 | 27150 |


| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | - | - | - | - | 515 | - | - | - | - | - | - |
| Germany, <br> Fed.Rep. | - | - | - | - | - | - | - | - | - | - | - |
| Ireland | 25200 | 16325 | 10164 | 11278 | 13072 | 12921 | 10950 | 13351 | 14840 | 12662 | 10237 |
| Netherlands | 2500 | 1868 | 1234 | 2088 | 366 | - | 64 | - | 353 | 13 |  |
| UK <br> (N.Ireland) | - | - | - | - | - | - | - | - | - | - | - |
| UK <br> (England + <br> Wales) | - | - | - | - | - | - | - | - | - | - | - |
| UK Scotland | - | - | - | - | - | - | - | - | 6 | - | - |
| Total landings | 27700 | 18193 | 11398 | 13366 | 13953 | 12921 | 11014 | 13351 | 15199 | 12675 | 10237 |
| Area misreported | 11200 | 7916 | 8448 | 1390 | 3873 | 3581 | 2813 | 2880 | 4353 | 5129 | 3103 |
| Unallocated |  |  |  |  |  |  |  |  | -353 | -13 |  |
| Discards | - | - | - | - | - | - | - | - | - | - | - |
| WG catch | 38900 | 26109 | 19846 | 14756 | 17826 | 16502 | 13827 | 16231 | 19193 | 17791 | 13340 |

Table 6.2.1.1 VIa(S) \& VIIb,c herring. Catch in numbers-at-age (winter rings) from 1970 to 2008.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 135 | 35114 | 26007 | 13243 | 3895 | 40181 | 2982 | 1667 | 1911 |
| 1971 | 883 | 6177 | 7038 | 10856 | 8826 | 3938 | 40553 | 2286 | 2160 |
| 1972 | 1001 | 28786 | 20534 | 6191 | 11145 | 10057 | 4243 | 47182 | 4305 |
| 1973 | 6423 | 40390 | 47389 | 16863 | 7432 | 12383 | 9191 | 1969 | 50980 |
| 1974 | 3374 | 29406 | 41116 | 44579 | 17857 | 8882 | 10901 | 10272 | 30549 |
| 1975 | 7360 | 41308 | 25117 | 29192 | 23718 | 10703 | 5909 | 9378 | 32029 |
| 1976 | 16613 | 29011 | 37512 | 26544 | 25317 | 15000 | 5208 | 3596 | 15703 |
| 1977 | 4485 | 44512 | 13396 | 17176 | 12209 | 9924 | 5534 | 1360 | 4150 |
| 1978 | 10170 | 40320 | 27079 | 13308 | 10685 | 5356 | 4270 | 3638 | 3324 |
| 1979 | 5919 | 50071 | 19161 | 19969 | 9349 | 8422 | 5443 | 4423 | 4090 |
| 1980 | 2856 | 40058 | 64946 | 25140 | 22126 | 7748 | 6946 | 4344 | 5334 |
| 1981 | 1620 | 22265 | 41794 | 31460 | 12812 | 12746 | 3461 | 2735 | 5220 |
| 1982 | 748 | 18136 | 17004 | 28220 | 18280 | 8121 | 4089 | 3249 | 2875 |
| 1983 | 1517 | 43688 | 49534 | 25316 | 31782 | 18320 | 6695 | 3329 | 4251 |
| 1984 | 2794 | 81481 | 28660 | 17854 | 7190 | 12836 | 5974 | 2008 | 4020 |
| 1985 | 9606 | 15143 | 67355 | 12756 | 11241 | 7638 | 9185 | 7587 | 2168 |
| 1986 | 918 | 27110 | 24818 | 66383 | 14644 | 7988 | 5696 | 5422 | 2127 |
| 1987 | 12149 | 44160 | 80213 | 41504 | 99222 | 15226 | 12639 | 6082 | 10187 |
| 1988 | 0 | 29135 | 46300 | 41008 | 23381 | 45692 | 6946 | 2482 | 1964 |
| 1989 | 2241 | 6919 | 78842 | 26149 | 21481 | 15008 | 24917 | 4213 | 3036 |
| 1990 | 878 | 24977 | 19500 | 151978 | 24362 | 20164 | 16314 | 8184 | 1130 |
| 1991 | 675 | 34437 | 27810 | 12420 | 100444 | 17921 | 14865 | 11311 | 7660 |
| 1992 | 2592 | 15519 | 42532 | 26839 | 12565 | 73307 | 8535 | 8203 | 6286 |
| 1993 | 191 | 20562 | 22666 | 41967 | 23379 | 13547 | 67265 | 7671 | 6013 |
| 1994 | 11709 | 56156 | 31225 | 16877 | 21772 | 13644 | 8597 | 31729 | 10093 |
| 1995 | 284 | 34471 | 35414 | 18617 | 19133 | 16081 | 5749 | 8585 | 14215 |
| 1996 | 4776 | 24424 | 69307 | 31128 | 9842 | 15314 | 8158 | 12463 | 6472 |
| 1997 | 7458 | 56329 | 25946 | 38742 | 14583 | 5977 | 8351 | 3418 | 4264 |
| 1998 | 7437 | 72777 | 80612 | 38326 | 30165 | 9138 | 5282 | 3434 | 2942 |
| 1999 | 2392 | 51254 | 61329 | 34901 | 10092 | 5887 | 1880 | 1086 | 949 |
| 2000 | 4101 | 34564 | 38925 | 30706 | 13345 | 2735 | 1464 | 690 | 1602 |
| 2001 | 2316 | 21717 | 21780 | 17533 | 18450 | 9953 | 1741 | 1027 | 508 |
| 2002 | 4058 | 32640 | 37749 | 18882 | 11623 | 10215 | 2747 | 1605 | 644 |
| 2003 | 1731 | 32819 | 28714 | 24189 | 9432 | 5176 | 2525 | 923 | 303 |
| 2004 | 1401 | 15122 | 32992 | 19720 | 9006 | 4924 | 1547 | 975 | 323 |
| 2005 | 209 | 28123 | 30896 | 26887 | 10774 | 5452 | 1348 | 858 | 243 |
| 2006 | 598 | 22036 | 36700 | 30581 | 21956 | 9080 | 2418 | 832 | 369 |
| 2007 | 76 | 24577 | 43958 | 23399 | 13738 | 5474 | 1825 | 231 | 131 |
| 2008 | 483 | 12265 | 19661 | 28483 | 11110 | 5989 | 2738 | 745 | 267 |

Table 6.2.1.2 VIa(S) \& VIIb,c herring. Percentage age composition (winter rings).

|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9 +}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |
| $\mathbf{1 9 9 4}$ | 6 | 28 | 15 | 8 | 11 | 7 | 4 | 16 | 5 |
| $\mathbf{1 9 9 5}$ | 0 | 23 | 23 | 12 | 13 | 11 | 4 | 6 | 9 |
| $\mathbf{1 9 9 6}$ | 3 | 13 | 38 | 17 | 5 | 8 | 4 | 7 | 4 |
| $\mathbf{1 9 9 7}$ | 5 | 34 | 16 | 23 | 9 | 4 | 5 | 2 | 3 |
| $\mathbf{1 9 9 8}$ | 3 | 29 | 32 | 15 | 12 | 4 | 2 | 1 | 1 |
| $\mathbf{1 9 9 9}$ | 1 | 30 | 36 | 21 | 6 | 3 | 1 | 1 | 1 |
| $\mathbf{2 0 0 0}$ | 3 | 27 | 30 | 24 | 10 | 2 | 1 | 1 | 1 |
| 2001 | 2 | 23 | 23 | 18 | 19 | 10 | 2 | 1 | 1 |
| 2002 | 3 | 27 | 31 | 16 | 10 | 9 | 2 | 1 | 1 |
| 2003 | 2 | 31 | 27 | 23 | 9 | 5 | 2 | 1 | 0 |
| 2004 | 2 | 18 | 38 | 23 | 10 | 6 | 2 | 1 | 0 |
| 2005 | 0 | 27 | 29 | 26 | 10 | 5 | 1 | 1 | 0 |
| 2006 | 0 | 18 | 29 | 25 | 18 | 7 | 2 | 1 | 0 |
| 2007 | 0 | 22 | 39 | 21 | 12 | 5 | 2 | 0 | 0 |
| 2008 | 1 | 15 | 24 | 35 | 14 | 7 | 3 | 1 | 0 |

Table 6.2.2.1 VIa(S) and VIIb,c herring. Sampling intensity of catches in 2008.

| ICES area | Year | Quarter | Landings (t) | No. Samples | No. aged | No. Measured | Aged/1000 t |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| VIaS official | 2008 | 1 | 4793 | 15 | 865 | 3536 | 180 |
| VIaS official | 2008 | 4 | 5077 | 17 | 939 | 3139 | 185 |
| VIIb | 2008 | 4 | 364 | 1 | 45 | 194 | 124 |
|  |  |  |  |  |  |  |  |
| Total North West |  |  | 10234 | 65 | 3653 | 13544 | 767 |

Table 6.2.2.2. VIa(S) and VIIb,c herring. Length distribution of Irish catches/quarter (thousands) 2008.

| Length cm | Quarter 1 <br> VIa South | Quarter 4 <br> VIIbc | Quarter 4 <br> VIa South |
| :---: | :---: | :---: | :---: |
| 20 |  |  |  |
| 20.5 |  |  |  |
| 21 |  |  |  |
| 21.5 |  |  | 28 |
| 22 | 9 |  | 46 |
| 22.5 | 63 |  | 139 |
| 23 | 134 |  | 232 |
| 23.5 | 233 |  | 335 |
| 24 | 699 | 10 | 669 |
| 24.5 | 1102 | 20 | 706 |
| 25 | 1684 | 10 | 1106 |
| 25.5 | 2275 | 40 | 1738 |
| 26 | 3153 | 69 | 2723 |
| 26.5 | 3878 | 276 | 3745 |
| 27 | 5320 | 375 | 4721 |
| 27.5 | 5410 | 375 | 4581 |
| 28 | 4156 | 385 | 4312 |
| 28.5 | 2194 | 237 | 2435 |
| 29 | 923 | 99 | 1069 |
| 29.5 | 287 | 30 | 353 |
| 30 | 54 |  | 130 |
| 30.5 | 27 |  | 37 |
| 31 | 9 |  | 46 |
| 31.5 | 27 |  |  |
| 32 |  |  | 9 |
| 32.5 |  |  |  |
| 33 | 18 |  |  |
| 33.5 | 9 |  |  |
| 34 |  |  |  |
| 34.5 |  |  |  |
| 35 |  |  |  |
| 35.5 |  |  |  |
| 36 | 9 |  |  |
| Nos./t | 31670 | 1926 | 29132 |

Table 6.3.1.1. VIa(S) \& VIIb,c herring. Time series of acoustic surveys since 1999. The 2008 survey is part of a new summer survey of the Malin Shelf stock complex.

| Winter rings | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |
| $\mathbf{0}$ | - | - | 5 | 0 | - | 0.09 | 1.28 | 0 | - | - |
| $\mathbf{1}$ | 18.99 | 10.71 | 22.69 | 35.7 | 10.28 |  | 7.83 | 1.6 | 0.3 | 12.28 |
| $\mathbf{2}$ | 104.77 | 60.88 | 52.33 | 14.05 | 26.26 | 3.9 | 56.91 | 6.9 | 3.5 | 83.33 |
| $\mathbf{3}$ | 32.53 | 48.96 | 6.41 | 24.23 | 30.02 | 62.35 | 93.51 | 86.7 | 59.8 | 64.85 |
| $\mathbf{4}$ | 11.34 | 25.57 | 6.47 | 14 | 11.08 | 54.93 | 109.87 | 57.5 | 21.9 | 38.02 |
| $\mathbf{5}$ | 1.65 | 9.43 | 2.63 | 5.79 | 2.94 | 80.07 | 100.8 | 27.9 | 11.7 | 22.04 |
| $\mathbf{6}$ | 0.94 | 2.35 | 1.94 | 5.7 | 0.64 | 47.14 | 56.54 | 16 | 6.35 | 28.67 |
| $\mathbf{7}$ | 0.3 | 1.28 | 0.12 | 5.06 | 0.94 | 13.81 | 21.16 | 4.8 | 1.86 | 9.03 |
| $\mathbf{8}$ | 0.17 | 0.43 | 0.24 | 2.73 | 0.3 | 11.77 | 24.64 | 4.8 | - | 4.99 |
| 9+ | 0.11 | 0.75 | 0.07 | 4.07 | 0.14 | - | 12.74 | 1.3 | - | 2.07 |
|  |  |  |  |  |  |  |  |  |  |  |
| Abundance (millions) | 170.8 | 160.36 | 97.9 | 111.33 | 82.6 | 274.06 | 485.29 | 202.9 | 105.41 | 266.85 |
| Total Biomass (t) | 23,762 | 21,048 | 11,062 | 8,867 | 10,300 | 41,700 | 71,253 | 27,770 | 14,222 | 44,611 |
| SSB (t) | 22,788 | 20,500 | 9,800 | 6,978 | 9,500 | 41,300 | 66,138 | 27,200 | 13,974 | 43,006 |
| CV\% | - | - | - | - | - | - | - | $49 \%$ | $44 \%$ | $34 \%$ |

Table 6.3.1.1. VIa(S) \& VIIb,c herring. Details of all acoustic surveys conducted on this stock.

| Year | Type | Biomass | SSB |
| :--- | :--- | :---: | :---: |
|  |  |  |  |
| 1994 | Feeding phase | - | 353,772 |
| 1995 | Feeding phase | 137,670 | 125,800 |
| 1996 | Feeding phase | 34,290 | 12,550 |
| 1997 | - | - | - |
| 1998 | - | - | - |
| 1999 | Autumn spawners | 23,762 | 22,788 |
| 2000 | Autumn spawners | 21,000 | 20,500 |
| 2001 | Autumn spawners | 11,100 | 9,800 |
| 2002 | Winter spawners | 8,900 | 7,200 |
| 2003 | Winter spawners | 10,300 | 9,500 |
| 2004 | Winter spawners | 41,700 | 41,399 |
| 2005 | Winter spawners | 71,253 | 66,138 |
| 2006 | Winter spawners | 27,770 | 27,200 |
| 2007 | Winter spawners | 14,222 | 13,974 |
| 2008 | Feeding phase | 44,611 | 43,006 |

Table 6.6.2.1. VIa(S) and VIIb,c herring VPA run with a terminal F value of 0.2

|  | RECRUITS 1r | SSB | LANDINGS | $\begin{gathered} \text { FBAR } \\ 3-6 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: |
| 1970 | 417093 | 140605 | 20306 | 0.1657 |
| 1971 | 840935 | 124426 | 15044 | 0.1476 |
| 1972 | 757077 | 131141 | 23474 | 0.1897 |
| 1973 | 552443 | 170952 | 36719 | 0.2697 |
| 1974 | 613024 | 101218 | 36589 | 0.4213 |
| 1975 | 429689 | 108757 | 38764 | 0.4 |
| 1976 | 721943 | 76291 | 32767 | 0.4532 |
| 1977 | 613201 | 86458 | 20567 | 0.2851 |
| 1978 | 1125083 | 81511 | 19715 | 0.2362 |
| 1979 | 1051379 | 116899 | 22608 | 0.2411 |
| 1980 | 570237 | 113116 | 30124 | 0.3453 |
| 1981 | 724411 | 116812 | 24922 | 0.2691 |
| 1982 | 751807 | 128324 | 19209 | 0.1943 |
| 1983 | 2457128 | 125443 | 32988 | 0.3157 |
| 1984 | 1021733 | 205755 | 27450 | 0.1789 |
| 1985 | 1291478 | 208768 | 23343 | 0.1508 |
| 1986 | 983832 | 245176 | 28785 | 0.161 |
| 1987 | 3345049 | 216039 | 48600 | 0.3086 |
| 1988 | 493834 | 326252 | 29100 | 0.244 |
| 1989 | 728143 | 243615 | 29210 | 0.1663 |
| 1990 | 820249 | 210562 | 43969 | 0.2408 |
| 1991 | 506336 | 179979 | 37700 | 0.2281 |
| 1992 | 418367 | 143988 | 31856 | 0.2626 |
| 1993 | 618832 | 123326 | 36763 | 0.3426 |
| 1994 | 805939 | 102494 | 33908 | 0.352 |
| 1995 | 459748 | 86792 | 27792 | 0.4549 |
| 1996 | 836771 | 64258 | 32534 | 0.5724 |
| 1997 | 828177 | 65604 | 27225 | 0.524 |
| 1998 | 533601 | 53327 | 38895 | 1.0014 |
| 1999 | 394353 | 45801 | 26109 | 0.6778 |
| 2000 | 452821 | 38580 | 19846 | 0.5119 |
| 2001 | 469261 | 36058 | 14756 | 0.6012 |
| 2002 | 599118 | 35335 | 17826 | 0.6524 |
| 2003 | 524671 | 42274 | 16502 | 0.5788 |
| 2004 | 583779 | 46840 | 13727 | 0.4955 |
| 2005 | 713483 | 50420 | 16231 | 0.4536 |
| 2006 | 500380 | 55201 | 19193 | 0.5741 |
| 2007 | 263972 | 54762 | 17791 | 0.3596 |
| 2008 | 700934* | 49554 | 13340 | 0.2797 |

*Geometric Mean 1970-2007

Table 6.6.2.2. VIa(S) and VIIbc herring VPA run using a terminal F or 0.4.

|  | RECRUITS 1-r | SSB | LANDINGS | FBAR 3-6 |
| :---: | :---: | :---: | :---: | :---: |
| 1970 | 417326 | 140862 | 20306 | 0.1654 |
| 1971 | 841453 | 124662 | 15044 | 0.1474 |
| 1972 | 757630 | 131386 | 23474 | 0.1895 |
| 1973 | 552932 | 171346 | 36719 | 0.2694 |
| 1974 | 613642 | 101419 | 36589 | 0.4207 |
| 1975 | 430274 | 108999 | 38764 | 0.3993 |
| 1976 | 722986 | 76484 | 32767 | 0.4522 |
| 1977 | 614217 | 86687 | 20567 | 0.2843 |
| 1978 | 1127264 | 81744 | 19715 | 0.2356 |
| 1979 | 1053710 | 117235 | 22608 | 0.2403 |
| 1980 | 571486 | 113486 | 30124 | 0.3442 |
| 1981 | 725825 | 117262 | 24922 | 0.268 |
| 1982 | 753320 | 128806 | 19209 | 0.1936 |
| 1983 | 2461691 | 125985 | 32988 | 0.3144 |
| 1984 | 1023586 | 206468 | 27450 | 0.1782 |
| 1985 | 1293324 | 209441 | 23343 | 0.1502 |
| 1986 | 984980 | 245905 | 28785 | 0.1604 |
| 1987 | 3348270 | 216737 | 48600 | 0.3077 |
| 1988 | 494187 | 327031 | 29100 | 0.2433 |
| 1989 | 728464 | 244192 | 29210 | 0.1659 |
| 1990 | 820393 | 211061 | 43969 | 0.2403 |
| 1991 | 506311 | 180341 | 37700 | 0.2277 |
| 1992 | 418286 | 144268 | 31856 | 0.2623 |
| 1993 | 618669 | 123544 | 36763 | 0.3424 |
| 1994 | 805553 | 102652 | 33908 | 0.3519 |
| 1995 | 459268 | 86814 | 27792 | 0.4549 |
| 1996 | 835238 | 64232 | 32534 | 0.5727 |
| 1997 | 825120 | 65506 | 27225 | 0.5247 |
| 1998 | 529962 | 53101 | 38895 | 1.0047 |
| 1999 | 389351 | 45422 | 26109 | 0.6828 |
| 2000 | 444350 | 38015 | 19846 | 0.5183 |
| 2001 | 451666 | 35225 | 14756 | 0.6152 |
| 2002 | 559187 | 33858 | 17826 | 0.6786 |
| 2003 | 465517 | 39217 | 16502 | 0.6183 |
| 2004 | 482152 | 41591 | 13727 | 0.5528 |
| 2005 | 528908 | 41693 | 16231 | 0.5357 |
| 2006 | 315965 | 40098 | 19193 | 0.7612 |
| 2007 | 146590 | 33273 | 17791 | 0.5633 |
| 2008 | 651814* | 24241 | 13340 | 0.5477 |

*Geometric mean 1970-2007.

Table 6.6.2.3 VIa(S) and VIIb,c herring VPA run using a terminal F or 0.5

|  | Recruitment 1r | SSB | Landings | FBAR $3-6 \mathrm{r}$ |
| ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |
| 1970 | 417413 | 140949 | 20306 | 0.1653 |
| 1971 | 841646 | 124741 | 15044 | 0.1473 |
| 1972 | 757833 | 131469 | 23474 | 0.1894 |
| 1973 | 553108 | 171482 | 36719 | 0.2693 |
| 1974 | 613860 | 101491 | 36589 | 0.4205 |
| 1975 | 430475 | 109085 | 38764 | 0.399 |
| 1976 | 723335 | 76554 | 32767 | 0.4519 |
| 1977 | 614550 | 86768 | 20567 | 0.284 |
| 1978 | 1127968 | 81825 | 19715 | 0.2354 |
| 1979 | 1054452 | 117349 | 22608 | 0.2401 |
| 1980 | 571879 | 113608 | 30124 | 0.3438 |
| 1981 | 726268 | 117410 | 24922 | 0.2677 |
| 1982 | 753791 | 128962 | 19209 | 0.1933 |
| 1983 | 2463112 | 126159 | 32988 | 0.314 |
| 1984 | 1024164 | 206694 | 27450 | 0.178 |
| 1985 | 1293903 | 209653 | 23343 | 0.1501 |
| 1986 | 985343 | 246133 | 28785 | 0.1602 |
| 1987 | 3349302 | 216956 | 48600 | 0.3074 |
| 1988 | 494302 | 327276 | 29100 | 0.243 |
| 1989 | 728577 | 244373 | 29210 | 0.1657 |
| 1990 | 820454 | 211219 | 43969 | 0.2402 |
| 1991 | 506319 | 180457 | 37700 | 0.2276 |
| 1992 | 418276 | 144359 | 31856 | 0.2622 |
| 1993 | 618643 | 123616 | 36763 | 0.3423 |
| 1994 | 805485 | 102706 | 33908 | 0.3519 |
| 1995 | 459177 | 86828 | 27792 | 0.4549 |
| 1996 | 834941 | 64232 | 32534 | 0.5727 |
| 1997 | 824524 | 65490 | 27225 | 0.5248 |
| 1998 | 529248 | 53058 | 38895 | 1.0053 |
| 1999 | 388366 | 45349 | 26109 | 0.6837 |
| 2000 | 442687 | 37906 | 19846 | 0.5196 |
| 2001 | 448252 | 35063 | 14756 | 0.618 |
| 2002 | 551525 | 33569 | 17826 | 0.6839 |
| 2003 | 454103 | 38625 | 16502 | 0.6267 |
| 2004 | 462124 | 40576 | 13727 | 0.5656 |
| 2005 | 491885 | 39988 | 16231 | 0.5554 |
| 2006 | 278765 | 37094 | 19193 | 0.8135 |
| 2007 | 28963 | 17791 | 0.6344 |  |
| 2008 | 19127 | 13340 | 0.6785 |  |
|  |  |  |  |  |

[^2]Table 6.6.2.4 VIa(S) and VIIb,c herring VPA run using a terminal F or 0.6

|  | RECRUITS 1-r | SSB | LANDINGS | FBAR 3-6 |
| :---: | :---: | :---: | :---: | :---: |
| 1970 | 417192 | 140712 | 20306 | 0.1656 |
| 1971 | 841260 | 124520 | 15044 | 0.1475 |
| 1972 | 757549 | 131244 | 23474 | 0.1896 |
| 1973 | 552984 | 171147 | 36719 | 0.2696 |
| 1974 | 613840 | 101339 | 36589 | 0.421 |
| 1975 | 430568 | 108924 | 38764 | 0.3996 |
| 1976 | 723667 | 76457 | 32767 | 0.4523 |
| 1977 | 614967 | 86709 | 20567 | 0.2842 |
| 1978 | 1128975 | 81811 | 19715 | 0.2354 |
| 1979 | 1055574 | 117408 | 22608 | 0.24 |
| 1980 | 572472 | 113720 | 30124 | 0.3434 |
| 1981 | 726921 | 117565 | 24922 | 0.2672 |
| 1982 | 754448 | 129164 | 19209 | 0.1929 |
| 1983 | 2464955 | 126394 | 32988 | 0.3135 |
| 1984 | 1024870 | 207006 | 27450 | 0.1777 |
| 1985 | 1294575 | 209946 | 23343 | 0.1498 |
| 1986 | 985745 | 246441 | 28785 | 0.16 |
| 1987 | 3350412 | 217245 | 48600 | 0.307 |
| 1988 | 494425 | 327581 | 29100 | 0.2428 |
| 1989 | 728696 | 244594 | 29210 | 0.1656 |
| 1990 | 820524 | 211405 | 43969 | 0.24 |
| 1991 | 506334 | 180589 | 37700 | 0.2274 |
| 1992 | 418275 | 144460 | 31856 | 0.2621 |
| 1993 | 618631 | 123696 | 36763 | 0.3422 |
| 1994 | 805447 | 102766 | 33908 | 0.3518 |
| 1995 | 459122 | 86846 | 27792 | 0.4549 |
| 1996 | 834752 | 64238 | 32534 | 0.5728 |
| 1997 | 824137 | 65484 | 27225 | 0.5249 |
| 1998 | 528782 | 53032 | 38895 | 1.0056 |
| 1999 | 387718 | 45303 | 26109 | 0.6843 |
| 2000 | 441591 | 37835 | 19846 | 0.5204 |
| 2001 | 446012 | 34956 | 14756 | 0.6198 |
| 2002 | 546548 | 33380 | 17826 | 0.6875 |
| 2003 | 446693 | 38239 | 16502 | 0.6323 |
| 2004 | 448975 | 39914 | 13727 | 0.5744 |
| 2005 | 467312 | 38873 | 16231 | 0.5691 |
| 2006 | 253905 | 35112 | 19193 | 0.852 |
| 2007 | 107398 | 26098 | 17791 | 0.692 |
| 2008 | 638235* | 15697 | 13340 | 0.8072 |

[^3]

Figure 6.1.2.1. VIa(S) \& VIIb,c herring. Working group estimate of catches from 1970-2008.
Northwest Herring Total Landings 2008


Figure 6.1.3.1. VIa(S) \& VIIb,c herring, Herring landings by statistical rectangle in VIaS and VIIbc in 2008.


Figure 6.2.1.1 VIa(S) \& Division VIIb,c herring. Mean standardised catch numbers at age standardised by year for the fishery. Numbers in thousands.


Depth contours 200-1000m
Figure 6.3.2.1. VIa(S) \& Division VIIb,c herring. Survey track for acoustic survey conducted in July 2008, in stock area. Conducted as part of Malin Shelf stock complex survey.


Figure 6.3.2.2. VIa(S) \& Division VIIb,c herring. Total NASC (nautical area scattering coefficient) for herring in acoustic survey conducted in July 2008, in stock area. Conducted as part to mixed traces.


Figure 6.4.1.1 VIa(S) \& Division VIIb,c herring. Mean Weights in the Catch (kg).


Figure 6.4.1.2 VIa(S) \& Division VIIb,c herring. Mean weights in the stock (kg).




Figure 6.6.2.1. VIa(S) and VIIb,c four separable VPA runs using values of $0.2,0.4,0.5$ and 0.6 for terminal F .


Figure 6.6.2.2. VIa(S) and VIIb,c herring - Residuals from three separable VPA runs using terminal $F$ values of $0.2,0.4,0.5$ and 0.6 . Red indicates positive residuals and white indicates negative


Figure 6.6.2.3. VIa(S) and VIIb,c herring. Comparison of four separable VPA runs of the current working group and the 2008 working group, using values of $0.2,0.4$ and 0.6 for terminal $F$.




Figure 6.6.2.4. VIa(S) and VIIb,c herring Retrospective assessment using $\mathrm{F}=\mathbf{0} \mathbf{0}$.




Figure 6.6.2.5. VIa(S) and VIIb,c herring Retrospective assessment using $\mathrm{F}=0.4$.




Figure 6.6.2.6 VIa(S) and VIIb,c herring Retrospective assessment using $\mathrm{F}=0.5$.




Figure 6.6.2.7. VIa(S) and VIIb,c herring Retrospective assessment using $\mathrm{F}=0.6$.

## 7 Irish Sea Herring [Division VIIa (North)]

### 7.1 The Fishery

### 7.1.1 Advice and management applicable to 2008 and 2009

The WG did not present the results of a final assessment to ACOM in 2008 due to the findings of preliminary data explorations. These explorations suggest that conflicting year effects are present in the acoustic and catch-at-age data, contributing to the poor model fit in the separable period. Though the exact level of the stock was unclear from the analysis, the trends from a two-stage biomass model suggested that the stock remains relatively stable. Acoustic and catch-at-age data both provided possible indications that a strong year class had entered the stock.

ACOM subsequently advised that a TAC of 4400 t , based on recent catches, be adopted for 2008. This advice was rejected in favour of a status quo TAC of 4800 t , partitioned as 3500 t to the UK and 1250 t to the Republic of Ireland.

### 7.1.2 The fishery in 2008

The catches reported from each country for the period 1986 to 2008 are given in Table 7.1.1, and total catches from 1961 to 2008 in Figure 7.1.1. Reported international landings in 2008 for the Irish Sea amounted to 4895 t with UK vessels acquiring extra quota through swaps with the Republic of Ireland. The majority of catches in 2008 were taken during the $3^{\text {rd }}$ quarter.

The 2008 VIIa(N) herring fishery opened in August, with the majority of catches taken during August and September by a pair of UK pair trawlers. September through to December saw activity of the Mourne fishery, limited to boats under 40 ft . This was the $3^{\text {rd }}$ year of recorded landings for this fishery. In 200823 vessels recorded landings of $\sim 153 \mathrm{t}$, the majority taken during October. The final take up of remaining TAC by the UK pair trawlers also took place during October.

### 7.1.3 Regulations and their effects

Closed areas for herring fishing in the Irish Sea along the east coast of Ireland and within 12 nautical miles of the west coast of Britain were maintained throughout the year. The traditional gillnet fishery on the Mourne herring, which has a derogation to fish within the Irish closed box, operated successfully again in 2008. The area to the east of the Isle of Man, encompassing the Douglas Bank spawning ground (described in ICES 2001, ACFM:10), was closed from $21^{\text {st }}$ September to $15^{\text {th }}$ November. Boats from the Republic of Ireland are not permitted to fish east of the Isle of Man.

The arrangement of closed areas in Division VIIa(N) prior to 1999 are discussed in detail in ICES (1996/ACFM:10) with a change to the closed area to the east of the Isle of Man being altered in 1999 (ICES 2001/ACFM:10). The closed areas consist of: all year juvenile closures along part of the east coast of Ireland, and the west coast of Scotland, England and Wales; spawning closures along the east coast of the Isle of Man from 21st September- 15th November, and along the east coast of Ireland all year round. The WG recommends that any alterations to the present closures be considered carefully, in the context of this report, to ensure protection for all components of this stock.

The TAC for VIIa(N) is partitioned as 3500 t to the UK and 1250 t to the Republic of Ireland.

### 7.1.4 Changes in fishing technology and fishing patterns

The fishery in area VIIa $(\mathrm{N})$ has not changed in recent years. A pair of UK pair trawlers takes the majority of catches during the $3^{\text {rd }}$ and $4^{\text {th }}$ quarters. A small local fishery continues to record landings on the traditional Mourne herring grounds during the $4^{\text {th }}$ quarter. This fishery has seen increasing catches of herring since 2006 with landings of $\sim 20 \mathrm{t}, \sim 33.5 \mathrm{t}$ in 2007 and $\sim 135 \mathrm{t}$ in 2008.

### 7.2 Biological Composition of the Catch

### 7.2.1 Catch in numbers

Catches in numbers-at-age are given in Table 7.2.1 for the years 1972 to 2008 and a graphical representation is given in Figure 7.2.1. The predominant year class in 2008 landings was the 2-ringers followed by the 1-ringers. The catch in numbers at length is given in Table 7.2.2 for 1993 to 2008.

### 7.2.2 Quality of catch and biological data

There are no estimates of discarding or slippage in the Irish Sea fisheries that target herring. Discarding however is not thought to be a feature of this fishery. Biological sampling remains high for this fishery with all data in 2008 arising from AFBI, Northern Ireland. It should be noted however that the majority of samples are taken from only one fishing unit, the pair of UK vessels operating in the Irish Sea. 19 samples were processed for 2008 with 18 from the $3^{\text {rd }}$ quarter fishery and 1 from the $4^{\text {th }}$ quarter. Further details of sampling are given in Table 7.2.3.

### 7.2.3 Acoustic surveys

The information on the time-series of acoustic surveys in the Irish Sea is given in Table 7.2.4. As in the last year's assessment, the SSB estimates from the survey are calculated using the (annually varying) maturity ogives from the commercial catch data.

The acoustic survey in 2008 was carried out over the period 27 August to 14 September. A survey design of stratified, systematic transects was employed, as in previous years (Figure 7.2.2.A). The bulk of the acoustic scatter attributed to pelagic fish was identified as sprat, which were abundant around the periphery of the Irish Sea and to the north west of the Isle of Man (Figure 7.2.2.B). However in recent years the ratio of sprat to herring has been seen to increase in favour of the 0 -group herring, a trend continued in 2008. 0-group herring were found to be most abundant to the east of the Isle of Man (Figure 7.2.3.B). 1+ herring targets were mostly distributed around the coasts of the Isle of Man (Figure 7.2.3.A). Further 1+ herring targets were found off the western Northern Irish coastline. In general, there are few samples on the age composition of the herring in the acoustic survey data. The survey followed the methods described in Armstrong et al., (ICES 2005 WD 23). Sampling intensity was high during the 2008 survey with 27 successful trawls completed. The length frequencies generated from these trawls highlights the spatial heterogeneous nature of herring age groups in the Irish Sea (Figure 7.2.4)

As in previous years, no herring schools were detected in the area immediately north of the Isle of Man, despite an abundance of early-stage larvae in this area in Novem-
ber (Figure 7.2.5). It is possible that spawning in this area only commences after the acoustic survey.

The estimate of herring SSB of 77172 t for 2008 is the highest estimate in the time series (Table 7.2.4). The approximate coefficient of variation (CV) of 0.23 is at the lower end of estimates associated with this survey. The biomass estimate of 106921 t for $1+$ ringers is the second highest estimate in the time series, whilst the approximate CV of 0.22 is also at the lower limits of this survey. The age-disaggregated acoustic estimates of the herring abundance, excluding 0-ring fish, is given in Table 7.2.5.

### 7.2.4 Larvae surveys

Northern Ireland undertook a herring larvae survey over the period $6^{\text {th }}$ to $17^{\text {th }}$ November 2008. The survey followed the methods and designs of previous surveys in the time-series (Annex 8). The production estimate for 2008 in the NE Irish Sea was a reduction on the previous year and below the time-series average (Table 7.2.6). As in previous years herring larvae were found to be most abundant to the south east and north east of the Isle of Man and less abundant in the western Irish Sea

Of note was the low occurrence of larvae in the area of the traditional Mourne spawning ground, where last year larvae had been caught. Signs of the expansion of a spawning component in this area in recent years are evident from the fishery operating here. As such larvae would be expected in the area. The low occurrence of larvae caught during the survey may therefore suggest a timing mis-match between larvae emergence and sampling.

### 7.3 Mean weight, maturity and natural mortality-at-age

Mean weights-at-age in the $3^{\text {rd }}$ quarter catches (for the whole time-series 1961 to present) have been used as estimates of stock weights at spawning time (Table 7.3.2). Maturity-at-age (in the catches) for each year (1961 to 2008) are given in Table 7.3.3. As in previous years, natural mortality per year was assumed to be 1.0 on 1-ringers, 0.3 for 2-ringers, 0.2 for 3-ringers and 0.1 for all older age classes (Annex 8). Mean weights-at-age have shown a general downward trend in the last 22 years.

### 7.4 Recruitment

An estimate of total abundance of 0-ringers and 1-ringers is provided by the Northern Ireland acoustic survey. However, there is evidence that a proportion of these are of Celtic Sea origin (Brophy and Danilowicz, 2002). Separation of the trawl catches of 0 -groups into autumn and winter spawning components, based on otolith microstructure and shape analysis was presented to the working group in 2008 by Beggs et al. (ICES 2008 WD4). It is hoped that repeating this procedure annually could result in a survey index of recruitment for the Irish Sea stock that could be used directly in the assessment. Such an index may also be of use in the Celtic Sea assessment, as it would provide an estimate of juveniles resident in the Irish Sea originating from this management area.

### 7.5 Stock Assessment

### 7.5.1 Data exploration and preliminary modelling

2008 data were added to the Northern Irish larvae series (NINEL), the Northern Irish acoustic survey (total biomass, SSB and age-structured indices) and the catch-at-age data derived from the landings.

During the 2008 WG, comparisons between total mortality rates estimated from the acoustic and catch-at-age data highlighted a divergence in estimates. The acoustic survey was shown to have higher estimates of total mortality due to year effects in the LnCatch ratios at ages 2-7. This divergence was considered to be associated with the variation in migration of herring enter the spawning area of the Irish Sea.

An exploratory SPALY assessment in 2009 confirmed that problems concerning residual patterns in the separable period of ICA remain. Therefore the results of the run are not presented, as they are not considered reliable for SSB and F during the separable period.

2008 acoustic survey estimates suggest that SSB is at higher levels than at anytime in the 14 year time-series, while 1-ringer+ biomass is also high. Numbers-at-age in the acoustic survey suggest the strong 2005 year class (1-ringers in 2007) is still present in the survey area as 2-ringers. This year class was also observed in the acoustic survey as a high abundance of 0 -groups in 2006. Microstructure analysis of the 0 -group otoliths classified approx. $90 \%$ of these juveniles in the eastern Irish Sea as "autumn" spawners Beggs et al., (ICES 2008 WD4). The 2005 strong year-class has now been tracked successfully over 3 years of the survey. Recruitment estimates of 0-group herring from the acoustic survey also remain high. The highest estimate of 0 -group herring in the time-series was observed in 2008, with the majority of biomass found in the eastern Irish Sea. This area has historically been associated with autumn spawning juveniles.

The strong 2005 year class was not as evident in the catch-at-age data from the 2008 fishery. Catch-at-age data did confirm the presence of relative high proportions of 5 and 6-ringers as also observed in the acoustic numbers-at-age.

Results of a microstructure analysis of 1-ringer+ fish were presented to the WG (ICES 2009 WD01). The study shows that "winter" spawners, of which the majority are thought to be of Celtic Sea origin, are present in the pre-spawning aggregations sampled in the Irish Sea during the acoustic survey. As previously suggested these fish are present in high proportions as 1-ringers but were also found as 2 and 3-ringers in varying proportions. The presence of these "winter" spawners has major implications for the estimates of 1-ringer+ biomass and SSB, as well as confounding traditional cohort type assessment methods, such as ICA.

### 7.5.2 Two-stage biomass model

In 2009 it was decided not to run the model in light of the SALY status of the stock assessment (see Annex 8).

### 7.5.3 Conclusion to explorations

The exploratory analysis to date suggests that the current configuration of ICA is unsuitable as an assessment method for the Irish Sea stock. Exploration of proportion-at-age data in 2007 suggests that conflicting year effects were present in the acoustic and catch-at-age data. These conflicting signals were contributing to the poor model fit in the separable period as shown by the large year residuals. In 2008 comparisons of the total mortality rates estimated from the acoustic and catch-at-age data suggested a conflicting signal with divergence in the estimates. Extensive mixing between fish of different seasonal origins during the acoustic survey introduces further residuals during the separable period. As a consequence of these effects recent estimates of SSB and F are unreliable, although trends in SSB and F during the converged period of the VPA are considered reliable.

Number-at-age data reveal the high portions of 1-and 2-ringers in the stock. The presence of the high numbers of 2-ringers currently in the stock can be substantiated by the strong recruitment event observed in 2005. This evidence suggests that a large year class is present in the fishery. The results of the microstructure study presented to the group suggest that Celtic Sea winter spawners are present in considerable numbers in the Irish Sea during the acoustic survey and fishery. The use of stock identity techniques to quantify their proportions at age is recommended.

### 7.5.4 Stock assessment

From the exploratory analysis it was considered that the current configuration of ICA is unsuitable for the assessment of this stock and therefore no runs are presented.

### 7.6 Stock and Catch Projection

### 7.6.1 Deterministic short-term predictions

The Working Group decided that there was no basis for undertaking short-term predictions of stock size.

### 7.6.2 Yield-per-recruit

The Working Group decided that there was no basis for yield-per-recruit analysis.

### 7.7 Medium-term predictions of stock size

The Working Group decided that there was no basis for undertaking medium-term projections of stock size.

### 7.8 Reference points

The estimation of $\mathbf{B}_{\text {pa }}\left(9500 \mathrm{t}\right.$ ) and $\mathbf{B}_{\lim }(6000 \mathrm{t})$ were not revisited this year. There were no new points to add to the discussions and deliberations presented in 2000 (ICES 2000/ACFM:12). There is no precautionary F value for this stock.

### 7.9 Quality of the assessment

An assessment of the stock was not conducted in 2009 in light of the exploratory analysis and the SALY (same advice as last year) status.

### 7.10 Management considerations

Given the historical landings from this stock and the knowledge that fishing pressure is mostly confined to one pair of UK vessels it can be assumed that fishing pressure and activity has not varied considerably in recent years. The catches have been close to TAC levels and the main fishing activity has not varied considerably as shown from landing data (Figure 7.1.1).

Acoustic data indicate that a strong year class may be present in the stock. Recent estimates of 0 -group herring biomass suggest continued strong recruitment. The growth of the Mourne fishery suggests that this stock or sub-component is under a state of expansion. The acoustic survey provides estimates of numbers-at-age, however the 1 to 3 -ringers in the area are a mixture of at least two adjacent stocks (Celtic Sea and VIIa(N))(Beggs et al., ICES 2009 WD01). Splitting of numbers-at-age into separate spawning components, based on otolith techniques could result in estimates more appropriate for the Irish Sea assessment.

Therefore the maintenance of catch levels at current levels $4800 t$, in the short-term, is considered precautionary.

A review of the model (ICA) configuration currently employed in the assessment of this stock is considered in light of the knowledge concerning the dynamics of this stock. The management and assessment of this stock is currently being evaluated under SGHERWAY.

### 7.11 Environment

### 7.11.1 Ecosystem Considerations

No additional information presented (see Annex 8)

### 7.11.2 Changes in Environment

No additional information presented (see Annex 8)

Table 7.1.1 Irish Sea Herring Division VIIa(N). Working group catch estimates in tonnes by country, 1987-2008. The total catch does not in all cases correspond to the official statistics and cannot be used for management purposes.

| Country | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ireland | 1200 | 2579 | 1430 | 1699 | 80 | 406 | 0 | 0 | 0 |
| UK | 3290 | 7593 | 3532 | 4613 | 4318 | 4864 | 4408 | 4828 | 5076 |
| Unallocated | 1333 | - | - | - | - | - | - | - | - |
| Total | 5823 | 10172 | 4962 | 6312 | 4398 | 5270 | 4408 | 4828 | 5076 |
|  |  |  |  |  |  |  |  |  |  |
| Country | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ |
| Ireland | 100 | 0 | 0 | 0 | 0 | 862 | 286 | 0 | 749 |
| UK | 5180 | 6651 | 4905 | 4127 | 2002 | 4599 | 2107 | 2399 | 1782 |
| Unallocated | 22 | - | - | - | - | - |  | - | - |
| Total | 5302 | 6651 | 4905 | 4127 | 2002 | 5461 | 2393 | 2399 | 2531 |
|  |  |  |  |  |  |  |  |  |  |
| Country | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |  |  |  |  |  |
| Ireland | 1153 | 581 | 0 | 0 |  |  |  |  |  |
| UK | 3234 | 3821 | 4629 | 4895 |  |  |  |  |  |
| Unallocated | - | - |  |  |  |  |  |  |  |
| Total | 4387 | 4402 | 4629 | 4895 |  |  |  |  |  |

Table 7.2.1 Irish Sea Herring Division VIIa(N). Catch-at-age (thousands) by year.

|  | AGe (rings) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1972 | 40640 | 46660 | 26950 | 13180 | 13750 | 6760 | 2660 | 1670 |
| 1973 | 42150 | 32740 | 38240 | 11490 | 6920 | 5070 | 2590 | 2600 |
| 1974 | 43250 | 109550 | 39750 | 24510 | 10650 | 4990 | 5150 | 1630 |
| 1975 | 33330 | 48240 | 39410 | 10840 | 7870 | 4210 | 2090 | 1640 |
| 1976 | 34740 | 56160 | 20780 | 15220 | 4580 | 2810 | 2420 | 1270 |
| 1977 | 30280 | 39040 | 22690 | 6750 | 4520 | 1460 | 910 | 1120 |
| 1978 | 15540 | 36950 | 13410 | 6780 | 1740 | 1340 | 670 | 350 |
| 1979 | 11770 | 38270 | 23490 | 4250 | 2200 | 1050 | 400 | 290 |
| 1980 | 5840 | 25760 | 19510 | 8520 | 1980 | 910 | 360 | 230 |
| 1981 | 5050 | 15790 | 3200 | 2790 | 2300 | 330 | 290 | 240 |
| 1982 | 5100 | 16030 | 5670 | 2150 | 330 | 1110 | 140 | 380 |
| 1983 | 1305 | 12162 | 5598 | 2820 | 445 | 484 | 255 | 59 |
| 1984 | 1168 | 8424 | 7237 | 3841 | 2221 | 380 | 229 | 479 |
| 1985 | 2429 | 10050 | 17336 | 13287 | 7206 | 2651 | 667 | 724 |
| 1986 | 4491 | 15266 | 7462 | 8550 | 4528 | 3198 | 1464 | 877 |
| 1987 | 2225 | 12981 | 6146 | 2998 | 4180 | 2777 | 2328 | 1671 |
| 1988 | 2607 | 21250 | 13343 | 7159 | 4610 | 5084 | 3232 | 4213 |
| 1989 | 1156 | 6385 | 12039 | 4708 | 1876 | 1255 | 1559 | 1956 |
| 1990 | 2313 | 12835 | 5726 | 9697 | 3598 | 1661 | 1042 | 1615 |
| 1991 | 1999 | 9754 | 6743 | 2833 | 5068 | 1493 | 719 | 815 |
| 1992 | 12145 | 6885 | 6744 | 6690 | 3256 | 5122 | 1036 | 392 |
| 1993 | 646 | 14636 | 3008 | 3017 | 2903 | 1606 | 2181 | 848 |
| 1994 | 1970 | 7002 | 12165 | 1826 | 2566 | 2104 | 1278 | 1991 |
| 1995 | 3204 | 21330 | 3391 | 5269 | 1199 | 1154 | 926 | 1452 |
| 1996 | 5335 | 17529 | 9761 | 1160 | 3603 | 780 | 961 | 1364 |
| 1997 | 9551 | 21387 | 7562 | 7341 | 1641 | 2281 | 840 | 1432 |
| 1998 | 3069 | 11879 | 3875 | 4450 | 6674 | 1030 | 2049 | 451 |
| 1999 | 1810 | 16929 | 5936 | 1566 | 1477 | 1989 | 444 | 622 |
| 2000 | 1221 | 3743 | 5873 | 2065 | 558 | 347 | 251 | 147 |
| 2001 | 2713 | 11473 | 7151 | 13050 | 3386 | 936 | 650 | 803 |
| 2002 | 179 | 9021 | 1894 | 1866 | 2395 | 953 | 474 | 343 |
| 2003 | 694 | 4694 | 3345 | 2559 | 882 | 2945 | 872 | 605 |
| 2004 | 3225 | 8833 | 5405 | 2161 | 623 | 213 | 673 | 127 |
| 2005 | 8692 | 13980 | 10555 | 3287 | 1422 | 415 | 292 | 368 |
| 2006 | 5669 | 15253 | 8198 | 6318 | 1325 | 605 | 262 | 246 |
| 2007 | 20290 | 18291 | 4980 | 1655 | 1062 | 325 | 122 | 111 |
| 2008 | 8939 | 18974 | 7487 | 2696 | 2082 | 1761 | 328 | 216 |

Table 7.2.2 Irish Sea Herring Division VIIa(N). Catch at length data 1993-2008. Numbers of fish in thousands. Table amended with 1990-1992 year-classes removed (see Annex 8).

| Length | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 14.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 15 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 15.5 |  |  |  |  |  |  | 10 |  |  |  |  |  |  |  | 16 |  |
| 16 |  |  | 21 | 21 | 17 |  | 19 | 12 | 9 |  |  |  |  | 2 |  |  |
| 16.5 |  |  | 55 | 51 | 94 |  | 53 | 49 | 27 |  |  | 13 | 1 | 44 | 33 |  |
| 17 |  | 84 | 139 | 127 | 281 | 26 | 97 | 67 | 53 |  |  | 25 | 39 | 140 | 69 | 3 |
| 17.5 |  | 59 | 148 | 200 | 525 | 30 | 82 | 97 | 105 |  |  | 84 | 117 | 211 | 286 | 11 |
| 18 |  | 69 | 300 | 173 | 1022 | 123 | 145 | 115 | 229 |  |  | 102 | 291 | 586 | 852 | 34 |
| 18.5 |  | 89 | 280 | 415 | 1066 | 206 | 135 | 134 | 240 | 36 |  | 114 | 521 | 726 | 2088 | 64 |
| 19 | 39 | 226 | 310 | 554 | 1720 | 317 | 234 | 164 | 385 | 18 |  | 203 | 758 | 895 | 2979 | 85 |
| 19.5 | 75 | 241 | 305 | 652 | 1263 | 277 | 82 | 97 | 439 | 0 | 29 | 269 | 933 | 1246 | 3527 | 108 |
| 20 | 75 | 253 | 326 | 749 | 1366 | 427 | 218 | 109 | 523 | 0 | 73 | 368 | 943 | 984 | 3516 | 100 |
| 20.5 | 57 | 270 | 404 | 867 | 1029 | 297 | 242 | 85 | 608 | 18 | 215 | 444 | 923 | 1443 | 2852 | 133 |
| 21 | 130 | 400 | 468 | 886 | 1510 | 522 | 449 | 115 | 1086 | 307 | 272 | 862 | 1256 | 1521 | 3451 | 192 |
| 21.5 | 263 | 308 | 782 | 1258 | 1192 | 549 | 362 | 138 | 1201 | 433 | 290 | 1007 | 1380 | 1621 | 2929 | 217 |
| 22 | 610 | 700 | 1509 | 1530 | 2607 | 1354 | 1261 | 289 | 1748 | 1750 | 463 | 1495 | 1361 | 2748 | 3821 | 271 |
| 22.5 | 1224 | 785 | 2541 | 2190 | 2482 | 1099 | 2305 | 418 | 1763 | 1949 | 600 | 2140 | 1448 | 3629 | 3503 | 229 |
| 23 | 2016 | 1035 | 4198 | 2362 | 3508 | 2493 | 4784 | 607 | 2670 | 2490 | 1158 | 2089 | 1035 | 4358 | 4196 | 322 |
| 23.5 | 2368 | 1473 | 4547 | 2917 | 3902 | 2041 | 4183 | 951 | 2254 | 1552 | 1380 | 2214 | 1256 | 2920 | 3697 | 26 |
| 24 | 2895 | 2126 | 4416 | 3649 | 4714 | 3695 | 4165 | 1436 | 3489 | 1029 | 1273 | 2054 | 1276 | 3679 | 3178 | 259 |
| 24.5 | 2616 | 2564 | 3391 | 4077 | 4138 | 2769 | 3397 | 1783 | 4098 | 758 | 1249 | 2269 | 1083 | 2431 | 2136 | 204 |
| 25 | 2207 | 3315 | 3100 | 4015 | 5031 | 2625 | 2620 | 2144 | 5566 | 776 | 1163 | 1749 | 1086 | 3438 | 1503 | 148 |
| 25.5 | 2198 | 3382 | 2358 | 3668 | 3971 | 2797 | 1817 | 1791 | 4785 | 1335 | 1211 | 1206 | 584 | 2198 | 952 | 114 |
| 26 | 2216 | 3480 | 2334 | 2480 | 3871 | 3115 | 1694 | 1349 | 3814 | 1570 | 1140 | 823 | 438 | 1714 | 643 | 78 |
| 26.5 | 2176 | 2617 | 1807 | 2177 | 2455 | 2641 | 1547 | 840 | 2243 | 1552 | 1573 | 587 | 203 | 605 | 330 | 42 |
| 27 | 2299 | 2391 | 1622 | 1949 | 1711 | 2992 | 1475 | 616 | 1489 | 776 | 1607 | 510 | 165 | 445 | 147 | 23 |
| 27.5 | 2047 | 1777 | 990 | 1267 | 1131 | 1747 | 867 | 479 | 644 | 433 | 1189 | 383 | 60 | 155 | 72 | 10 |
| 28 | 1538 | 1294 | 834 | 906 | 638 | 1235 | 276 | 212 | 496 | 162 | 726 | 198 | 45 | 104 | 33 | 12 |
| 28.5 | 944 | 900 | 123 | 564 | 440 | 170 | 169 | 58 | 179 | 108 | 569 | 51 | 18 | 9 | 26 | 1 |
| 29 | 473 | 417 | 248 | 210 | 280 | 111 | 61 | 42 | 10 | 36 | 163 |  | 12 | 46 |  |  |
| 29.5 | 160 | 165 | 56 | 79 | 59 | 92 |  | 12 | 0 | 36 | 129 |  |  |  | 7 |  |
| 30 | 83 | 9 | 40 | 32 | 8 | 84 |  | 6 | 9 |  | 43 |  |  |  |  |  |
| 30.5 | 15 | 27 | 5 | 0 | 5 | 3 |  |  |  |  | 43 |  |  |  |  |  |
| 31 | 4 |  | 1 | 2 |  |  |  |  |  |  | 43 |  |  |  |  |  |
| 31.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 32 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 32.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 33 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 33.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 34 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 7.2.3 Irish Sea Herring Division VIIa(N). Sampling intensity of commercial landings in 2008.

| Quarter | Country | Landings (T) | No. SAMPLES | No. FISH MEASURED | No. FISH AGED |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Ireland | 0 | - | - | - |
|  | UK (N. Ireland) | 0.095 | 0 | 0 | 0 |
|  | UK (Isle of Man) | 0 | - | - | - |
|  | UK (Scotland) | 0 | - | - | - |
|  | UK (England \& Wales) | 0 | - | - | - |
| 2 | Ireland | 0 | - | - | - |
|  | UK (N. Ireland) | 0 | - | - | - |
|  | UK (Isle of Man) | * | - | - | - |
|  | UK (Scotland) | 0 | - | - | - |
|  | UK (England \& Wales) | 0 | - | - | - |
| 3 | Ireland | 0 | - | - | - |
|  | UK (N. Ireland) | 4131 | 18 | 2790 | 888 |
|  | UK (Isle of Man) | * | - | - | - |
|  | UK (Scotland) | 0 | - | - | - |
|  | UK (England \& Wales) | 0 | - | - | - |
| 4 | Ireland | 0 | - | - | - |
|  | UK (N. Ireland) | 764 | 1 | 135 | 50 |
|  | UK (Isle of Man) | * | - | - | - |
|  | UK (Scotland) | 0 | - | - | - |
|  | UK (England \& Wales) | 0 | - | - | - |

[^4]Table 7.2.4 Irish Sea Herring Division VIIa(N). Summary of acoustic survey information for the period 1989-2008. Small clupeoids include sprat and 0-ring herring unless otherwise stated. CVs are approximate. Biomass in t . All surveys carried out at 38 kHz except December 1996, which was at 120 kHz .

| Year | AreA | DATES | HERRING BIOMASS | CV | HERRING BIOMASS | CV | SMALL CLUPEOI DS | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | (1+years) |  | (SSB) |  | biomass |  |
| 1989 | Douglas <br> Bank | 25/09-26/09 |  |  | 18,000 | - | - | - |
| 1990 | Douglas Bank | 26/09-27/09 |  |  | 26,600 | - | - | - |
| 1991 | W. Irish Sea | 26/07-8/08 | 12,760 | 0.23 |  |  | 66,000 ${ }^{1}$ | 0.20 |
| 1992 | W. Irish Sea + IOM E. coast | 20/07-31/07 | 17,490 | 0.19 |  |  | 43,200 | 0.25 |
| $1994$ | Area VIIa(N) | 28/08-8/09 | 31,400 | 0.36 | 25,133 | - | 68,600 | 0.10 |
|  | Douglas Bank | 22/09-26/09 |  |  | 28,200 | - | - | - |
| 1995 | Area VIIa(N) | 11/09-22/09 | 38,400 | 0.29 | 20,167 | - | 348,600 | 0.13 |
|  | Douglas Bank | 10/10-11/10 |  | - | 9,840 | - | - | - |
|  | Douglas Bank | 23/10-24/10 |  |  | 1,750 | 0.51 | - | - |
| 1996 | Area VIIa(N) | 2/09-12/09 | 24,500 | 0.25 | 21,426 | 0.25 | -2 | - |
| 1997 | Area VIIa(N)reduced | 8/09-12/09 | 20,100 | 0.28 | 10,702 | 0.35 | 46,600 | 0.20 |
| 1998 | Area VIIa(N) | 8/09-14/09 | 14,500 | 0.20 | 9,157 | 0.18 | 228,000 | 0.11 |
| 1999 | Area VIIa(N) | 6/09-17/09 | 31,600 | 0.59 | 21,040 | 0.75 | 272,200 | 0.10 |
| 2000 | Area VIIa(N) | 11/09-21/09 | 40,200 | 0.26 | 33,144 | 0.32 | 234,700 | 0.11 |
| 2001 | Area VIIa(N) | 10/09-18/09 | 35,400 | 0.40 | 13,647 | 0.42 | 299,700 | 0.08 |
| 2002 | Area VIIa(N) | 9/09-20/09 | 41,400 | 0.56 | 25,102 | 0.83 | 413,900 | 0.09 |
| 2003 | Area VIIa(N) | 7/09-20/09 | 49,500 | 0.22 | 24,390 | 0.24 | 265,900 | 0.10 |
| 2004 | Area VIIa(N) | $\begin{aligned} & \text { 6/09-10/09, } \\ & \text { 15/09-16/09, } \\ & 28 / 09-29 / 09 \end{aligned}$ | 34,437 | 0.41 | 21,593 | 0.41 | 281,000 | 0.07 |
| 2005 | Area VIIa(N) | 29/08-14/09 | 36,866 | 0.37 | 31,445 | 0.42 | 141,900 | 0.10 |
| 2006 | Area VIIa(N) | 30/08-9/09 | 33,136 | 0.24 | 16,332 | 0.22 | 143,200 | 0.09 |
| 2007 | Area VIIa(N) | 29/08-13/09 | 120,878 | 0.53 | 51,819 | 0.42 | 204,700 | 0.09 |
| 2008 | Area VIIa(N) | 27/08-14/09 | 106,921 | 0.22 | 77,172 | 0.23 | 252,300 | 0.12 |

[^5]Table 7.2.5 Irish Sea Herring Division VIIa(N). Age-disaggregated acoustic estimates (thousands) of herring abundance from the Northern Ireland surveys in September (ACAGE).

| AGE <br> (RINGS) | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1994 | 66.8 | 68.3 | 73.5 | 11.9 | 9.3 | 7.6 | 3.9 | 10.1 |
| 1995 | 319.1 | 82.3 | 11.9 | 29.2 | 4.6 | 3.5 | 4.9 | 6.9 |
| 1996 | 11.3 | 42.4 | 67.5 | 9 | 26.5 | 4.2 | 5.9 | 5.8 |
| 1997 | 134.1 | 50 | 14.8 | 11 | 7.8 | 4.6 | 0.6 | 1.9 |
| 1998 | 110.4 | 27.3 | 8.1 | 9.3 | 6.5 | 1.8 | 2.3 | 0.8 |
| 1999 | 157.8 | 77.7 | 34 | 5.1 | 10.3 | 13.5 | 1.6 | 6.3 |
| 2000 | 78.5 | 103.4 | 105.3 | 27.5 | 8.1 | 5.4 | 4.9 | 2.4 |
| 2001 | 387.6 | 93.4 | 10.1 | 17.5 | 7.7 | 1.4 | 0.6 | 2.2 |
| 2002 | 391 | 71.9 | 31.7 | 24.8 | 31.3 | 14.8 | 2.8 | 4.5 |
| 2003 | 349.2 | 220 | 32 | 4.7 | 3.9 | 4.1 | 1 | 0.9 |
| 2004 | 241 | 115.5 | 29.6 | 15.4 | 2.1 | 2.3 | 0.2 | 0.2 |
| 2005 | 94.3 | 109.9 | 97.1 | 17 | 8 | 0.8 | 0.6 | 5.8 |
| 2006 | 374.7 | 96.6 | 15.6 | 10.0 | 0.5 | 0.4 | 0.5 | 0.5 |
| 2007 | 1316.7 | 251.3 | 46.6 | 21.1 | 20.8 | 1.2 | 0.7 | 0.6 |
| 2008 | 475.7 | 452.4 | 114.2 | 39.1 | 26.4 | 17.1 | 4.3 | 0.6 |

Table 7.2.6 Irish Sea Herring Division VIIa(N). Larval production (10 ${ }^{11}$ ) indices for the Manx component. Table amended with Douglas Bank time series removed (see Annex 8).

| YEAR |  |  |  | Northeast IRISH SEA |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  | Isle of Man |  | Northern Ireland |  |  |

SE = Standard Error *2005 Index value amended

Table 7.3.2 Irish Sea Herring Division VIIa(N). Mean weights-at-age in the catch.

| Year | Weights-at-age (g) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age (rings) |  |  |  |  |  |  |  |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1985 | 87 | 125 | 157 | 186 | 202 | 209 | 222 | 258 |
| 1986 | 68 | 143 | 167 | 188 | 215 | 229 | 239 | 254 |
| 1987 | 58 | 130 | 160 | 175 | 194 | 210 | 218 | 229 |
| 1988 | 70 | 124 | 160 | 170 | 180 | 198 | 212 | 232 |
| 1989 | 81 | 128 | 155 | 174 | 184 | 195 | 205 | 218 |
| 1990 | 77 | 135 | 163 | 175 | 188 | 196 | 207 | 217 |
| 1991 | 70 | 121 | 153 | 167 | 180 | 189 | 195 | 214 |
| 1992 | 61 | 111 | 136 | 151 | 159 | 171 | 179 | 191 |
| 1993 | 88 | 126 | 157 | 171 | 183 | 191 | 198 | 214 |
| 1994 | 73 | 126 | 154 | 174 | 181 | 190 | 203 | 214 |
| 1995 | 72 | 120 | 147 | 168 | 180 | 185 | 197 | 212 |
| 1996 | 67 | 116 | 148 | 162 | 177 | 199 | 200 | 214 |
| 1997 | 64 | 118 | 146 | 165 | 176 | 188 | 204 | 216 |
| 1998 | 80 | 123 | 148 | 163 | 181 | 177 | 188 | 222 |
| 1999 | 69 | 120 | 145 | 167 | 176 | 188 | 190 | 210 |
| 2000 | 64 | 120 | 148 | 168 | 188 | 204 | 200 | 213 |
| 2001 | 67 | 106 | 139 | 156 | 168 | 185 | 198 | 205 |
| 2002 | 85 | 113 | 144 | 167 | 180 | 184 | 191 | 217 |
| 2003* | 81 | 116 | 136 | 160 | 167 | 172 | 186 | 199 |
| 2004 | 73 | 107 | 130 | 157 | 165 | 187 | 200 | 205 |
| 2005 | 67 | 103 | 136 | 156 | 166 | 180 | 191 | 209 |
| 2006 | 64 | 105 | 131 | 149 | 164 | 177 | 184 | 211 |
| 2007 | 67 | 112 | 135 | 158 | 173 | 183 | 199 | 227 |
| 2008 | 71 | 110 | 135 | 153 | 156 | 182 | 196 | 206 |

* Average for the preceding five years

Table 7.3.3 Irish Sea Herring Division VIIa(N). Maturity ogive (maturity in the catch).

| Year | AGe (rings) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1961 | 0.00 | 0.22 | 0.63 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1962 | 0.00 | 0.24 | 0.83 | 0.92 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1963 | 0.00 | 0.34 | 0.88 | 0.89 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1964 | 0.00 | 0.53 | 0.81 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1965 | 0.00 | 0.61 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1966 | 0.00 | 0.47 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1967 | 0.02 | 0.37 | 0.75 | 0.83 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1968 | 0.00 | 0.88 | 0.94 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1969 | 0.00 | 0.71 | 0.92 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1970 | 0.02 | 0.92 | 0.94 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1971 | 0.15 | 0.87 | 0.97 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1972 | 0.11 | 0.88 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1973 | 0.12 | 0.77 | 0.89 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1974 | 0.36 | 0.99 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1975 | 0.40 | 0.99 | 1.00 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1976 | 0.07 | 0.96 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1977 | 0.03 | 0.92 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1978 | 0.04 | 0.81 | 0.88 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1979 | 0.00 | 0.84 | 0.81 | 0.78 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1980 | 0.20 | 0.88 | 0.95 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1981 | 0.19 | 0.89 | 0.90 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1982 | 0.10 | 0.80 | 0.89 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1983 | 0.02 | 0.73 | 0.88 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1984 | 0.00 | 0.69 | 0.83 | 0.93 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1985 | 0.14 | 0.62 | 0.71 | 0.88 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1986 | 0.31 | 0.73 | 0.66 | 0.81 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1987 | 0.00 | 0.85 | 0.91 | 0.87 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1988 | 0.00 | 0.90 | 0.96 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1989 | 0.07 | 0.63 | 0.93 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1990 | 0.06 | 0.66 | 0.90 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1991 | 0.04 | 0.30 | 0.74 | 0.82 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1992 | 0.28 | 0.48 | 0.72 | 0.81 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1993 | 0.00 | 0.46 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1994 | 0.19 | 0.68 | 0.99 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1995 | 0.10 | 0.86 | 0.94 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1996 | 0.02 | 0.60 | 0.96 | 0.83 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.04 | 0.82 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1998 | 0.30 | 0.83 | 0.97 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1999 | 0.02 | 0.84 | 0.95 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2000 | 0.14 | 0.79 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2001 | 0.15 | 0.54 | 0.88 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2002 | 0.02 | 0.92 | 0.95 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2003* | 0.11 | 0.76 | 0.95 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2004 | 0.11 | 1.00 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.20 | 0.97 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2006 | 0.19 | 0.89 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2007 | 0.16 | 0.94 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2008 | 0.16 | 0.84 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

[^6]

Figure 7.1.1
Irish Sea herring VIIa(N). Landings of herring from VIIa(N) from 1961 to 2008.


Figure 7.2.1
Irish Sea herring VIIa(N). Landings (catch-at-age) of herring from VIIa(N) from 1961 to 2008.


Figure 7.2.2 Irish Sea herring VIIa(N). (A) Transects, stratum boundaries and trawl positions for the 2008 acoustic survey; (B) Density distribution of sprats (size of ellipses is proportional to square root of the fish density ( t n.mile ${ }^{-2}$ ) per 15 -minute interval). Maximum density was 800 t n.mile ${ }^{-2}$.


Figure 7.2.3 Irish Sea herring VIIa(N). (A) Density distribution of 1-ring and older herring (size of ellipses is proportional to square root of the fish density ( $\mathbf{t}$ n.mile ${ }^{-2}$ ) per 15-minute interval). Maximum density was 2670 t n.mile ${ }^{-2}$. (B) Density distribution of 0 -ring herring. Maximum density was 450 t n.mile ${ }^{-2}$. Note: same scaling of ellipse sizes on above figures.


Figure 7.2.4 Irish Sea herring VIIa(N). Percentage length compositions of herring in each trawl sample in the September 2008 acoustic survey.


Figure 7.2.5
Irish Sea herring VIIa(N). Estimates of larval herring abundance in the Northern Irish Sea, $6^{\text {th }}$ to $17^{\text {th }}$ November 2008. (maximum abundance $=71.01$ per m²).

## 8 Sprat in the North Sea

### 8.1 The Fishery

### 8.1.1 ACFM Advice Applicable to 2008 and 2009

There have never been any explicit management objectives for this stock. The TAC set for 2008 was 175000 t . A mid-year revise did not change the TAC in 2008. For 2008, the by-catch quota of herring (EU fleet) was set at 18806 t . For 2009 a preliminary TAC of 170000 t is set and a revised mid-year advice is expected. For 2009, the bycatch quota of herring (EU fleet) was set at 15985 t .

## Catches in 2008

Catch statistics for 1996-2008 for sprat in the North Sea by area and country are presented in Table 8.1.1. Catch data prior to 1996 are considered unreliable. In 1996 total landings were 137.000 t and have since been in the range of 61.000 t (2008) to 208.000 t (2005). As in previous years sprat from the fjords of western Norway are not included in the catches for the North Sea, due to uncertainties in stock identity. Annual catches of Norwegian fjord sprat have ranged between $400 t(2004)$ and $3300 t(1996,1999)$ in this period. Total catches for the North Sea in 2008 were 61083 t , the lowest for the entire time series. The Danish catches represent more than $95 \%$ of the total catches. The Norwegian sprat fishery caught 1266 t of sprat.

The catches by year, quarter, and area show the same picture as last year, with the largest amount taken in IVb and in IVc. Only small catches were landed in the first two quarters in 2008 (Table 8.1.2). Quarterly and annual distribution of catches per rectangle for Subarea IV show a fishery located in the mid-southern North Sea in the first and second quarter, while the central-eastern areas are targeted in the second half of the year (Figures 8.1.1a-d and Figure 8.1.2).

### 8.1.2 Regulations and their effects

The Norwegian vessels are not allowed to fish in the Norwegian zone until the quota in the EU-zone has been taken. They are not allowed to fish in the 2 nd and 3rd quarters in the EU and the Norwegian zone. There is a maximum vessel quota of 800 t . A herring by-catch of up to $10 \%$ in biomass is allowed in Norwegian sprat catches. A by-catch of up to $20 \%$ in biomass of herring is allowed in the Danish sprat catches. Sprat cannot be fished without by-catches of herring except in years with high sprat abundance or low herring recruitment. Management of this stock should consider management advice given for herring in Subarea IV, Division VIId, and Division IIIa. A decrease in recruitment for the North Sea herring autumn spawners and a probable high incoming sprat year class may potentially result in a fishery for sprat with less by-catch of herring.

Most sprat catches are taken in an industrial fishery where catches are limited by herring by-catch restrictions.

In 2007 a new quota regulation (IOK) for the Danish vessels was implemented and realized from 2008 and onwards. The regulation gives quotas to the vessel, but these can be traded or sold. A large number of small vessels have been taken out of the fishery and their quotas sold to larger vessels. Today the Danish fleet is therefore dominated by large vessels.

### 8.1.3 Changes in fishing technology and fishing patterns

No changes in fishing technology and fishing patterns of importance for the sprat fisheries in the North Sea have been reported.

### 8.2 Biological composition of the catch

Only data on by-catch from the Danish fishery were available to the Working Group (Table 8.2.1). The Danish sprat fishery has recently been conducted with a by-catch of herring. The total amount of herring caught as by-catch in the sprat fishery has mainly been less than $10 \%$ but increased to $11 \%$ in 2008.

The Danish biological sampling from 1996 and onwards is considered reliable due to the changes in the Danish sampling scheme and the estimated quarterly landings at age in numbers for the period are presented in Table 8.2.2. In 2008 the one-year old sprat contributed $44 \%$ of the total landings, which is the second lowest value since 1996 (2004: 35\%, all other years: 51-96\%). 2-year olds contributed in 2008 with $41 \%$ of the total landing, leaving $13 \%$ of the contribution to $0-3$ - and 4 -year olds.

Mean-weight-at-age (g) in the landings in 2008 was lower than the 2007 values for all year classes (Table 8.2.3). But the lower values are more in accordance with the years before 2007.

Denmark, Norway and UK-Scotland provided age data of commercial landings in 2008 for all quarters fished (Table 8.2.4). These data were used to raise the landings data from the North Sea. The landings by UK-England were minor and unsampled. The sampling level (no. per 1000 t ) in 2008 was similar to 2007 considering number of samples (0.4) and number aged (2008: 16, 2007: 18), but decreased considering the number measured. In 200840 sprat per 1000 t were measured compared to 57 per 1000 t in 2007. In Denmark the provisions in the EU regulation 1639/2001 and the amendment 1581/2004 have been implemented. This provision requires 1 sample per 2000 t landed. This sampling level is lower than the guidelines ( 1 sample per 1000 t ) previously used by the HAWG but as the main fishery was carried out in a limited area and a limited season, the recommended sampling level can be regarded as adequate.

### 8.3 Fishery Independent Information

### 8.3.1 IBTS (February)

The calculation of this index can be found in the stock annex.
Sprat of age 1 and 2 were found in the south-east, with the highest concentrations in the more central parts of the distribution area (Figure 8.3.1a-c) and Division IVc.

### 8.3.2 Acoustic Survey

The sprat was in 2008 almost exclusively found in the eastern and southern parts of the North Sea, with highest abundances mainly in the central southern part (Figure 8.3.2). Total abundance was estimated to 25125 million individuals and total biomass 271000 t which is a reduction by more than $20 \%$ in terms of biomass when compared to last year and the lowest estimate for the period 2003-2008 (ICES CM 2009/LCR:02). In 2008, as in most recent years the majority of the stock consists of mature sprat. In 2007 roughly $1 / 3$ of the sprat biomass was immature fish. The estimated strength of the 1-year-olds in 2008 (the 2007 year class) is the lowest since 2002. The sprat stock is dominated by 1 - and 2 -year old fish representing more than $95 \%$ of the biomass.

|  | ABUNDANCE (millions) |  |  |  |  | $\begin{aligned} & \text { BIOMASS } \\ & (1000 \mathrm{t}) \end{aligned}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year/Age | 0 | 1 | 2 | 3+ | sum | 0 | 1 | 2 | $3+$ | sum |
| 2008 | 0 | 17165 | 7410 | 549 | 25125 | 0.0 | 160.6 | 101.2 | 8.6 | 270.4 |
| 2007* | 0 | 37250 | 5513 | 1869 | 44631 | 0.0 | 258.0 | 66.2 | 28.5 | 352.7 |
| 2006* | 0 | 21862 | 19916 | 760 | 42537 | 0.0 | 158.9 | 265.2 | 11.8 | 435.9 |
| 2005* | 0 | 69798 | 2526 | 350 | 72674 | 0.0 | 474.6 | 32.8 | 5.9 | 513.3 |
| 2004* | 17401 | 28940 | 5312 | 367 | 52019 | 19.3 | 266.6 | 73.3 | 6.3 | 365.5 |
| 2003* | 0 | 25294 | 3983 | 338 | 29615 | 0.0 | 198.4 | 61.1 | 6.0 | 265.5 |
| 2002 | 0 | 15769 | 3687 | 207 | 19664 | 0.0 | 166.8 | 55.1 | 3.7 | 225.6 |
| 2001 | 0 | 12639 | 1812 | 110 | 14561 | 0.0 | 96.5 | 23.5 | 1.8 | 121.8 |
| 2000 | 0 | 11569 | 6407 | 180 | 18156 | 0.0 | 100.4 | 92.4 | 2.8 | 195.6 |
| 1999 | 0 | 353 | 5 | 0 | 358 | 0.0 | 3.3 | 0.0 | 0.0 | 3.3 |
| 1998 | 17 | 5365 | 960 | 37 | 6379 | 0.1 | 48.2 | 14.1 | 0.8 | 63.2 |

*Re-calculated by the means of FishFrame (ICES 2009/LRC:02)

### 8.3.3 Survey indices

The time series of the different survey indices for all ages and 1-year-old sprat are shown in Figure 8.3.3 and 8.3.4.

The survey indices for North Sea sprat is often strongly influenced by a few large hauls. A quantification of the importance of each haul to the index has been made by estimating the cumulative contribution of each hauls. The cumulative index ranks the 300-450 individual haul contributions to the IBTS Q1 sprat age 1 survey index (Figure 8.3.5). Individual hauls for each year are sorted by size and aggregated to calculate a cumulative distribution. For all years in the IBTS survey the largest 10 hauls contribute to $35-85 \%$ of the survey index. In the 2009 IBTS Q1 the largest haul contributes approximately $30 \%$ to the index. In exceptional years more than $50 \%$ of the index was driven by a single haul.

Estimates of the distribution of the IBTS Q1 indices are available from the ICES DATRAS database, based on a resampling from the original individual haul data ("bootstrapping"). Confidence intervals based on these resampled estimates (Figure 8.3.6) are extremely broad. The upper confidence limit ranges from $30 \%$ to $4600 \%$ greater than the value of the index estimated by ICES, with a median value of $250 \%$. The lower confidence limit ranges from $20 \%$ to $90 \%$ less than the value estimated by ICES, with a median value of $40 \%$. HAWG therefore concludes that the uncertainties in the value of this index are extremely broad, and dominate the dynamics of the index itself. There are no combinations of years in this time series where it is possible to say that there is a statistically significant difference in the estimated abundance of sprat.

## Management stocks

North Sea sprat is considered as an independent stock. This management approach has been tested by including IBTS survey data from the subdivisions VIId and IIIa (Figure 8.3.7a-c) for comparison of the CPUE for each statrec at which data are available. Data from subdivision VIId have been sampled during the French (2007) and Dutch (2008 and 2009) IBTS surveys. The North Sea management stock is framed by
the red line. No distinct separation is obvious between North Sea sprat and sprat in VIId, whereas IIIa sprat and North Sea sprat show a lesser overlap.

### 8.4 Mean weights-at-age and maturity-at-age

Data on maturity by age, mean weight- and length-at-age during the 2008 summer acoustic survey are presented in the PGIPS report (ICES 2009/LRC:02).

### 8.5 Recruitment

The IBTS (February) 1-group index is used as a recruitment index for this stock.
The 2005 index of 1-group (2004 year class) was the highest for the time series until this year (see Table 8.3.1). The high level of the 1-group in 2005 was seen in most samples and not only confined to a few single hauls. This year class was abundant as 3 -group in 2007, and in 2008 it was above, but near the average for the 4 -group. In 2009 the incoming 1-group (2008 year class) is estimated to be the highest for the whole time series, both in absolute and relative terms, but this estimate should be considered as preliminary. The index is also dominated by a few large hauls (see also 8.3.3).

### 8.6 Assessment of sprat in the North Sea

Previous exploratory assessments of this stock have been performed using the CSA method (ICES HAWG 2008). Generally, this method has given very poor results: the results are highly sensitive to the ratio of the survey catchabilities, $s$, and this parameter cannot be firmly estimated, either internally within the model or from external information. Similarly, the uncertainty associated with the corresponding estimates of biomass are extremely large, ranging over more than an order of magnitude - it is therefore not possible to say, with this method, whether the stock size of this population is less than one million tonnes, or greater than 10 million tonnes.

This method is therefore clearly inappropriate for the task at hand, and the results meaningless in an advice content. The decision was therefore made not to perform or report any such runs this year.

This stock will be the subject of a benchmark assessment in September 2009, as part of the WKSHORT workshop.

### 8.7 North Sea Sprat Forecasts

In previous years, a catch prediction for the assessment year was provided on the basis of a linear regression of catch (as estimated by landings) versus the IBTS sprat index summed over all age groups. Following issues raised in HAWG 2008, and subsequent discussions during this meeting, HAWG concluded that such an approach had no scientific merit, for three reasons.

Firstly, the fishery is limited in some years by the TAC, whilst in other years it is not. The fishery is also opportunistic in nature, and the intensity of the North Sea sprat fishery is greatly influenced by the quotas set for other preferred industrial species, such as sandeel. It is therefore not reasonable to expect a consistent relationship between stock size and catch.

Secondly the traditional catch regression also includes all years from 1984 onwards however, there are concerns about the quality of landing information prior to 1996 and this information was not used in previous exploratory assessments.

Finally, as highlighted elsewhere in this report (Section 8.3.3), the uncertainty associated with the IBTS estimate of stock-size is extremely high, due to the fact that it is driven by a few large catches. Making a catch prediction based on this index value would therefore seem unwise.

Given these problems, HAWG concluded that this approach has no meaning in an advice context. No forecast for this stock is presented.

### 8.8 Quality of the Assessment

No quantitative assessment is presented for this stock.
Uncertainties in the survey indices make the current understanding of the dynamics of this stock extremely poor. HAWG recommends that the detailed study of improved or alternative assessment methods (e.g. length based assessment) and the use of additional information sources (e.g. acoustic surveys, catch per unit effort) are required in order improve our level of understanding and ability to adequately manage this stock.

### 8.9 State of the Stock

Precautionary reference points have not been defined for this stock and the available information is inadequate to estimate the absolute stock size. Relative trends in abundance from indices suggest that the stock has dropped appreciably from its mid 2000s high, and is now in the lower quartile of observed values. (Figures 8.3.3-4).

### 8.10 Management Considerations

### 8.11 Ecosystem Considerations

Multispecies investigations have demonstrated that sprat is one of the important prey species in the North Sea ecosystem. Many of the plankton-feeding fish have recruited poorly in recent years (e.g. herring, sandeel, Norway pout). The implications of the environmental change for sprat and the influence of the sprat fishery for other fish species and sea birds, are at present unknown.

The zooplankton community structure that is sustaining the sprat stocks appears to be changing, and there has been a long-term decrease in total zooplankton abundance in the northern North Sea (Reid et al., 2003; Beaugrand, 2003; ICES, 2006). However, sprat is mainly distributed in the southern North Sea where these trends have not been observed (ICES, 2006).

### 8.12 Changes in the environment

Temperatures in this area have been increasing over the last number of decades. It is considered that this may have implications for sprat, although it is not possible to quantify either the magnitude or direction of such changes.

Table 8.1.1. North Sea sprat. Catches (' 000 t) 1996-2008. See ICES CM 2006/ACFM:20
for earlier catch data. Catch in fjords of western Norway excluded.
(Data provided by Working Group members except where indicated). These figures do not in all cases correspond to the official statistics and cannot be used for management purposes.
The IVb catches for 2000-2007 divided by IVbW and IVE can be found in ICES CM 2008/ACOM:02

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Division IVa |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 0.3 |  |  | 0.7 |  | 0.1 | 1.1 |  | * |  | * | 0.8 | * |
| Sweden |  |  |  |  |  | 0.1 |  |  |  |  |  |  |  |
| Total |  |  |  | 0.7 |  | 0.2 | 1.1 |  | * |  | * | 0.8 | * |
| Division Ivb |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 76.5 | 93.1 | 119.3 | 160.3 | 162.9 | 143.9 | 126.1 | 152.9 | 175.9 | 204.0 | 79.5 | 55.5 | 51.4 |
| Norway | 52.8 | 3.1 | 15.3 | 13.1 | 0.9 | 5.9 | * |  | 0.1 |  | 0.8 | 3.7 | 1.3 |
| Sweden | 0.5 |  | 1.7 | 2.1 |  | 1.4 |  |  |  | * |  |  |  |
| UK(Scotland) |  |  |  | 1.4 |  |  |  |  |  |  |  | 0.1 |  |
| Total |  | 96.2 | 136.3 | 176.9 | 163.8 | 151.2 | 126.1 | 152.9 | 176.0 | 204.1 | 80.3 | 59.3 | 52.7 |
| Division IVc |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 3.9 | 5.7 | 11.8 | 3.3 | 28.2 | 13.1 | 14.8 | 22.3 | 16.8 | 2.0 | 23.8 | 20.6 | 8.1 |
| Netherlands |  |  |  | 0.2 |  |  |  |  |  |  |  |  |  |
| Norway |  | 0.1 | 16.0 | 5.7 | 1.8 | 3.6 |  |  |  |  | 9.0 | 2.9 |  |
| UK(Engl. \&Wales) | 2.6 | 1.4 | 0.2 | 1.6 | 2.0 | 2.0 | 1.6 | 1.3 | 1.5 | 1.6 | 0.5 | 0.3 | * |
| UK(Scotland) |  |  |  |  |  |  |  |  |  |  |  |  | 0.2 |
| Total |  | 7.2 | 28.0 | 10.8 | 32.0 | 18.7 | 16.4 | 23.6 | 18.3 | 3.6 | 33.4 | 23.8 | 8.4 |
| Total North Sea |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Denmark | 80.7 | 98.8 | 131.1 | 164.3 | 191.1 | 157.1 | 142.0 | 175.2 | 192.7 | 206.0 | 103.4 | 76.8 | 59.6 |
| Netherlands |  |  |  | 0.2 |  |  |  |  |  |  |  |  |  |
| Norway | 52.8 | 3.2 | 31.3 | 18.8 | 2.7 | 9.5 | * |  | 0.1 |  | 9.8 | 6.7 | 1.3 |
| Sweden | 0.5 |  | 1.7 | 2.1 |  | 1.5 |  |  |  | * |  |  |  |
| UK(Engl. \&Wales) | 2.6 | 1.4 | 0.2 | 1.6 | 2.0 | 2.0 | 1.6 | 1.3 | 1.5 | 1.6 | 0.5 | 0.3 | * |
| UK(Scotland) |  |  |  | 1.4 |  |  |  |  |  |  |  | 0.1 | 0.2 |
| Total | 136.6 | 103.4 | 164.3 | 188.4 | 195.9 | 170.2 | 143.6 | 176.5 | 194.3 | 207.7 | 113.7 | 83.8 | 61.1 |

* < 50 t

Table 8.1.2. North Sea sprat. Catches (tonnes) by quarter. Catches in fjords
of Western Norway excluded. Data for 1996-1999 in ICES CM 2007/ACFM :11
The IVb catches for 2000-2007 divided by IVbW and IVE can be found in ICES CM 2008/ACOM:02

| Year | Quarter |  | Area |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | IVaW | IVaE | IVb | IVc |  |
| 2000 | 1 |  |  | 18126 | 28063 | 46189 |
|  | 2 |  |  | 1722 | 45 | 1767 |
|  | 3 |  |  | 131306 | 1216 | 132522 |
|  | 4 |  |  | 12680 | 2718 | 15398 |
|  | Total |  |  | 163834 | 32042 | 195876 |
| 2001 | 1 | 115 |  | 40903 | 9716 | 50734 |
|  | 2 |  |  | 1071 |  | 1071 |
|  | 3 |  |  | 44174 | 481 | 44655 |
|  | 4 | 79 |  | 65102 | 8538 | 73719 |
|  | Total | 194 |  | 151249 | 18735 | 170177 |
| 2002 | 1 | 1136 |  | 2182 | 2790 | 6108 |
|  | 2 |  |  | 435 | 93 | 528 |
|  | 3 |  |  | 70504 | 647 | 71151 |
|  | 4 |  |  | 52942 | 12911 | 65853 |
|  | Total | 1136 |  | 126063 | 16441 | 143640 |
| 2003 | 1 |  |  | 11458 | 7727 | 19185 |
|  | 2 |  |  | 625 | 26 | 652 |
|  | 3 |  |  | 56207 | 165 | 56372 |
|  | 4 |  |  | 84629 | 15651 | 100280 |
|  | Total |  |  | 152919 | 23570 | 176489 |
| 2004 | 1 |  |  | 827 | 1831 | 2657 |
|  | 2 | 7 |  | 260 | 16 | 283 |
|  | 3 |  |  | 54161 | 496 | 54657 |
|  | 4 |  |  | 120685 | 15937 | 136622 |
|  | Total | 7 |  | 175932 | 18280 | 194219 |
| 2005 | 1 |  |  | 11538 | 2457 | 13995 |
|  | 2 |  |  | 2515 | 123 | 2638 |
|  | 3 |  |  | 107530 |  | 107530 |
|  | 4 |  |  | 82474 | 1033 | 83507 |
|  | Total |  |  | 204057 | 3613 | 207670 |
| 2006 | 1 | 25 | 22 | 13713 | 33534 | 47294 |
|  | 2 |  |  | 190 | 8 | 198 |
|  | 3 |  |  | 40051 | 8 | 40059 |
|  | 4 | 2 |  | 26579 | 77 | 26658 |
|  | Total | 27 | 22 | 80533 | 33627 | 114209 |
| 2007 | 1 |  |  | 582 | 247 | 829 |
|  | 2 |  |  | 241 | 3 | 244 |
|  | 3 |  |  | 16603 |  | 16603 |
|  | 4 | 769 |  | 41850 | 23531 | 66150 |
|  | Total | 769 |  | 59276 | 23781 | 83826 |
| 2008 | 1 |  |  | 2872 | 43 | 2915 |
|  | 2 |  |  | 52 | * | 52 |
|  | 3 |  |  | 21787 |  | 21787 |
|  | 4 |  |  | 27994 | 8334 | 36329 |
|  | Total |  |  | 52706 | 8377 | 61083 |

* < 0.5 t

Any inconsistency in total catches is due to rounding errors.

Table 8.2.1. North Sea sprat. Species composition in the Danish sprat fishery in tonnes and percentage of the total catch. Data is reported for 1998-2008

|  | Year | Sprat | Herring | Horse mack. | Whiting | Haddock | Mackerel | Cod | Sandeel | Other | Total |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Tonnes | 1998 | 129315 | 11817 | 573 | 673 | 6 | 220 | 11 | 2174 | 1188 | 145978 |  |
| Tonnes | 1999 | 157003 | 7256 | 413 | 1088 | 62 | 321 | 7 | 4972 | 635 | 171757 |  |
| Tonnes | 2000 | 188463 | 11662 | 3239 | 2107 | 66 | 766 | 4 | 423 | 1911 | 208641 |  |
| Tonnes | 2001 | 136443 | 13953 | 67 | 1700 | 223 | 312 | 4 | 17020 | 1142 | 170862 |  |
| Tonnes | 2002 | 140568 | 16644 | 2078 | 2537 | 27 | 715 | 0 | 4102 | 800 | 167471 |  |
| Tonnes | 2003 | 172456 | 10244 | 718 | 1106 | 15 | 799 | 11 | 5357 | 3509 | 194214 |  |
| Tonnes | 2004 | 179944 | 10144 | 474 | 334 |  | 4351 | 3 | 3836 | 1821 | 200906 |  |
| Tonnes | 2005 | 201331 | 21035 | 2477 | 545 | 4 | 1009 | 16 | 6859 | 974 | 234 | 250 |
| Tonnes | 2006 | 103236 | 8983 | 577 | 343 | 25 | 905 | 4 | 5384 | 576 | 120033 |  |
| Tonnes | 2007 | 74734 | 6596 | 168 | 900 | 6 | 126 | 18 | 6 | 253 | 82807 |  |
| Tonnes | 2008 | 61093 | 7928 | 26 | 380 | 10 | 367 | 0 | 23 | 1735 | 71563 |  |
| Percent | 1998 | 88.6 | 8.1 | 0.4 | 0.5 | 0.0 | 0.2 | 0.0 | 1.5 | 0.8 | 100.0 |  |
| Percent | 1999 | 91.4 | 4.2 | 0.2 | 0.6 | 0.0 | 0.2 | 0.0 | 2.9 | 0.4 | 100.0 |  |
| Percent | 2000 | 90.3 | 5.6 | 1.6 | 1.0 | 0.0 | 0.4 | 0.0 | 0.2 | 0.9 | 100.0 |  |
| Percent | 2001 | 79.9 | 8.2 | 0.0 | 1.0 | 0.1 | 0.2 | 0.0 | 10.0 | 0.7 | 100.0 |  |
| Percent | 2002 | 83.9 | 9.9 | 1.2 | 1.5 | 0.0 | 0.4 | 0.0 | 2.4 | 0.5 | 100.0 |  |
| Percent | 2003 | 88.8 | 5.3 | 0.4 | 0.6 | 0.0 | 0.4 | 0.0 | 2.8 | 1.8 | 100.0 |  |
| Percent | 2004 | 89.6 | 5.0 | 0.2 | 0.2 | 0.0 | 2.2 | 0.0 | 1.9 | 0.9 | 100.0 |  |
| Percent | 2005 | 85.9 | 9.0 | 1.1 | 0.2 | 0.0 | 0.4 | 0.0 | 2.9 | 0.4 | 100.0 |  |
| Percent | 2006 | 86.0 | 7.5 | 0.5 | 0.3 | 0.0 | 0.8 | 0.0 | 4.5 | 0.5 | 100.0 |  |
| Percent | 2007 | 90.3 | 8.0 | 0.2 | 1.1 | 0.0 | 0.2 | 0.0 | 0.0 | 0.3 | 100.0 |  |
| Percent | 2008 | 85.4 | 11.1 | 0.0 | 0.5 | 0.0 | 0.5 | 0.0 | 0.0 | 2.4 | 100.0 |  |

Table 8.2.2 North Sea Sprat. Catch in numbers (millions) by quarter and by age 1996-2008.

| Year | Quarter | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0 | 1 | 2 | 3 | 4 | 5+ | Total |
| 1996 | 1 |  | 524.7 | 4615.4 | 2621.9 | 316.4 | 11.3 | 8090 |
|  | 2 |  | 1.9 | 241.5 | 32.7 | 15.5 | 0.3 | 292 |
|  | 3 |  | 400.5 | 100.7 | 22.9 | 0.3 |  | 524 |
|  | 4 |  | 1190.7 | 1069.0 | 339.6 | 5.6 |  | 2605 |
|  | Total |  | 2117.8 | 6026.6 | 3017.1 | 337.8 | 11.6 | 11511 |
| 1997 | 1 |  | 74.4 | 314.0 | 229.2 | 55.3 | 2.5 | 675 |
|  | 2 |  | 11.3 | 47.8 | 34.9 | 8.4 | 0.4 | 103 |
|  | 3 |  | 1991.9 |  |  |  |  | 1992 |
|  | 4 | 127.6 | 3597.2 | 996.2 | 117.8 | 58.1 |  | 4897 |
|  | Total | 127.6 | 5674.8 | 1358.1 | 381.9 | 121.8 | 2.8 | 7667 |
| 1998 | 1 |  | 683.2 | 537.2 | 18.3 | 0.1 |  | 1239 |
|  | 2 |  | 70.9 | 55.3 | 1.8 |  |  | 128 |
|  | 3 | 74.2 | 3356.6 | 693.3 |  |  |  | 4124 |
|  | 4 | 772.4 | 4822.4 | 2295.1 | 483.5 | 39.5 |  | 8413 |
|  | Total | 846.6 | 8933.1 | 3580.9 | 503.6 | 39.6 |  | 13904 |
| 1999 | 1 |  | 728.1 | 2226.0 | 554.2 | 86.6 | 9.2 | 3604 |
|  | 2 |  | 38.6 | 58.4 | 18.1 | 2.6 |  | 118 |
|  | 3 |  | 12919.0 | 38.9 |  |  |  | 12958 |
|  | 4 | 105.0 | 2143.2 | 211.5 |  |  |  | 2460 |
|  | Total | 105.0 | 15828.9 | 2534.8 | 572.3 | 89.2 | 9.2 | 19139 |
| 2000 | 1 |  | 559.2 | 3177.3 | 797.5 | 247.5 | 72.0 | 4854 |
|  | 2 |  | 6.8 | 107.4 | 60.1 | 12.8 | 0.5 | 188 |
|  | 3 |  | 9928.9 | 1111.9 | 77.8 |  |  | 11119 |
|  | 4 |  | 1153.7 | 129.2 | 9.0 |  |  | 1292 |
|  | Total |  | 11648.7 | 4525.8 | 944.4 | 260.3 | 72.6 | 17452 |
| 2001 | 1 |  | 746.3 | 3197.7 | 1321.9 | 22.2 |  | 5288 |
|  | 2 |  | 15.9 | 66.2 | 26.1 |  |  | 108 |
|  | 3 | 0.4 | 3338.8 | 299.9 |  |  |  | 3639 |
|  | 4 | 1205.0 | 4178.7 | 1224.6 | 261.9 |  |  | 6870 |
|  | Total | 1205.4 | 8279.8 | 4788.4 | 1609.9 | 22.2 |  | 15906 |
| 2002 | 1 |  | 104.7 | 400.3 | 30.2 | 11.2 |  | 546 |
|  | 2 |  | 13.7 | 27.9 | 2.4 | 0.6 |  | 45 |
|  | 3 | 40.9 | 5745.6 | 582.1 | 42.3 | 4.1 |  | 6415 |
|  | 4 | 415.0 | 4578.0 | 626.2 | 119.8 | 3.1 |  | 5742 |
|  | Total | 455.9 | 10441.9 | 1636.5 | 194.8 | 19.0 |  | 12748 |
| 2003 | 1 |  | 1953.9 | 1218.9 | 85.3 | 11.3 |  | 3269 |
|  | 2 |  | 41.8 | 46.3 | 4.7 | 0.6 |  | 93 |
|  | 3 | 1.1 | 3481.3 | 772.0 | 42.9 |  |  | 4297 |
|  | 4 | 539.3 | 7051.8 | 1115.1 | 93.8 | 36.5 | 21.9 | 8858 |
|  | Total | 540.4 | 12528.7 | 3152.3 | 226.6 | 48.4 | 21.9 | 16518 |
| 2004 | 1 |  | 16.5 | 214.0 | 26.3 | 1.6 | 0.6 | 259 |
|  | 2 |  | 22.1 | 14.9 | 3.0 | 0.1 |  | 40 |
|  | 3 | 210.0 | 3661.9 | 558.2 | 31.4 |  |  | 4462 |
|  | 4 | 15674.4 | 5582.8 | 632.1 | 59.2 |  |  | 21949 |
|  | Total | 15884.4 | 9283.2 | 1419.2 | 119.8 | 1.8 | 0.6 | 26709 |
| 2005 | 1 |  | 2476.5 | 268.5 | 13.8 | 2.2 |  | 2761 |
|  | 2 |  | 499.6 | 23.4 | 4.3 | 4.9 |  | 532 |
|  | 3 |  | 11920.2 | 192.3 | 7.6 |  |  | 12120 |
|  | 4 | 302.5 | 7467.9 | 191.1 |  |  |  | 7962 |
|  | Total | 302.5 | 22364.3 | 675.3 | 25.7 | 7.0 |  | 23375 |
| 2006 | 1 |  | 1559.2 | 5119.1 | 95.7 | 2.3 |  | 6776 |
|  | 2 |  | 5.8 | 21.5 | 0.2 |  |  | 27 |
|  | 3 |  | 3077.8 | 625.0 | 129.1 |  |  | 3832 |
|  | 4 |  | 2048.5 | 416.0 | 85.9 |  |  | 2550 |
|  | Total |  | 6691.2 | 6181.6 | 310.8 | 2.3 |  | 13186 |
| 2007 | 1 |  | 12.1 | 57.4 | 17.3 |  |  | 87 |
|  | 2 |  | 3.9 | 18.5 | 5.6 |  |  | 28 |
|  | 3 |  | 1025.3 | 194.5 | 17.7 | 25.3 |  | 1263 |
|  | 4 | 858.6 | 4047.6 | 1066.0 | 150.9 |  |  | 6123 |
|  | Total | 858.6 | 5088.8 | 1336.5 | 191.4 | 25.3 |  | 7501 |
| 2008 | 1 |  | 356.0 | 170.9 | 8.4 | 1.0 |  | 536 |
|  | 2 |  | 7.8 | 2.7 | 0.1 |  |  | 11 |
|  | 3 | 1.7 | 444.3 | 1225.8 | 189.9 | 29.3 |  | 1891 |
|  | 4 | 486.3 | 1812.5 | 1032.8 | 147.5 | 13.9 |  | 3493 |
|  | Total | 488.0 | 2620.5 | 2432.2 | 345.9 | 44.2 |  | 5931 |

Table 8.2.3 North Sea Sprat. Mean weight (g) by quarter and by age for 1996-2008.

| Year | Quarter | Age |  |  |  |  |  | SOP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0 | 1 | 2 | 3 | 4 | $5+$ | Tonnes |
| 1996 | 1 |  | 3.9 | 9.3 | 14.9 | 15.3 | 16.1 | 88807 |
|  | 2 |  | 6.9 | 8.4 | 11.6 | 20.0 | 15.2 | 2735 |
|  | 3 |  | 11.6 | 14.2 | 18.2 | 21.5 |  | 6501 |
|  | 4 |  | 12.1 | 15.9 | 17.2 | 20.5 |  | 37359 |
| Weighted mean |  |  | 10.0 | 10.5 | 15.1 | 15.6 | 16.0 | 135401 |
| 1997 | 1 |  | 8.0 | 10.0 | 15.0 | 17.0 | 19.0 | 8161 |
|  | 2 |  | 8.0 | 10.0 | 15.0 | 17.0 | 19.0 | 1243 |
|  | 3 |  | 14.2 |  |  |  |  | 28285 |
|  | 4 | 3.7 | 11.9 | 16.4 | 19.1 | 19.6 |  | 63083 |
| Weighted mean |  | 3.7 | 12.7 | 14.7 | 16.3 | 18.2 | 19.0 | 100772 |
| 1998 | 1 |  | 5.6 | 6.0 | 8.7 | 15.0 |  | 7232 |
|  | 2 |  | 5.6 | 6.0 | 8.3 |  |  | 743 |
|  | 3 | 3.7 | 14.7 | 15.3 |  |  |  | 60149 |
|  | 4 | 4.1 | 10.6 | 13.8 | 16.3 | 14.6 |  | 94173 |
| Weighted mean |  | 4.0 | 11.7 | 12.8 | 16.0 | 14.7 |  | 162297 |
| 1999 | 1 |  | 3.3 | 8.7 | 12.5 | 14.4 | 16.3 | 30168 |
|  | 2 |  | 3.1 | 10.1 | 13.6 | 15.4 |  | 993 |
|  | 3 |  | 10.0 | 18.3 |  |  |  | 129383 |
|  | 4 | 4.4 | 11.0 | 14.4 |  |  |  | 27126 |
| Weighted mean |  | 4.4 | 9.8 | 9.4 | 12.5 | 14.4 | 16.3 | 187670 |
| 2000 | 1 |  | 4.2 | 10.1 | 10.7 | 10.2 | 10.5 | 46192 |
|  | 2 |  | 3.3 | 9.0 | 10.2 | 12.8 | 10.5 | 1767 |
|  | 3 |  | 11.9 | 11.9 | 11.0 |  |  | 132563 |
|  | 4 |  | 11.9 | 11.9 | 11.0 |  |  | 15403 |
| Weighted mean |  |  | 11.6 | 10.6 | 10.7 | 10.3 | 10.5 | 195925 |
| 2001 | 1 |  | 3.3 | 9.7 | 12.9 | 16.5 |  | 50794 |
|  | 2 |  | 3.3 | 10.3 | 12.9 |  |  | 1071 |
|  | 3 | 4.0 | 12.0 | 15.3 |  |  |  | 44656 |
|  | 4 | 3.8 | 11.6 | 12.6 | 19.1 |  |  | 73444 |
| Weighted mean |  | 3.8 | 11.0 | 10.8 | 13.9 | 16.5 |  | 169965 |
| 2002 | 1 |  | 7.0 | 12.0 | 14.0 | 13.0 |  | 61057 |
|  | 2 |  | 5.3 | 11.2 | 12.5 | 12.4 |  | 4231 |
|  | 3 | 2.0 | 10.9 | 15.0 | 15.0 | 24.0 |  | 721732 |
|  | 4 | 3.9 | 12.0 | 15.0 | 15.7 | 24.0 |  | 679018 |
| Weighted mean |  | 3.7 | 11.2 | 13.4 | 14.9 | 14.8 |  | 1466038 |
| 2003 | 1 |  | 3.6 | 9.4 | 11.0 | 15.0 |  | 19599 |
|  | 2 |  | 3.1 | 9.9 | 11.0 | 15.0 |  | 648 |
|  | 3 | 3.0 | 13.0 | 16.0 | 13.0 |  |  | 58169 |
|  | 4 | 4.6 | 10.8 | 14.8 | 16.9 | 15.0 | 18.0 | 97670 |
| Weighted mean |  | 4.6 | 10.3 | 12.9 | 13.8 | 15.0 | 18.0 | 176085 |
| 2004 | 1 |  | 3.6 | 10.3 | 13.8 | 16.6 | 16.1 | 2663 |
|  | 2 |  | 6.0 | 8.5 | 7.3 | 10.2 |  | 282 |
|  | 3 | 4.5 | 11.9 | 17.0 | 20.0 |  |  | 54639 |
|  | 4 | 4.0 | 11.4 | 14.6 | 18.3 |  |  | 136653 |
| Weighted mean |  | 4.0 | 11.0 | 10.9 | 14.5 | 16.8 | 16.1 | 194238 |
| 2005 | 1 |  | 4.6 | 8.9 | 12.1 | 16.0 |  | 13995 |
|  | 2 |  | 4.8 | 6.5 | 9.8 | 10.0 |  | 2641 |
|  | 3 |  | 8.9 | 9.9 | 18.6 |  |  | 107531 |
|  | 4 | 4.1 | 10.7 | 12.0 |  |  |  | 83515 |
| Weighted mean |  | 4.1 | 8.9 | 10.0 | 13.6 | 11.8 |  | 207682 |
| 2006 | 1 |  | 4.3 | 7.7 | 9.6 | 13.0 |  | 47293 |
|  | 2 |  | 3.7 | 8.1 | 11.2 |  |  | 198 |
|  | 3 |  | 9.8 | 12.5 | 16.1 |  |  | 40053 |
|  | 4 |  | 9.8 | 12.5 | 16.1 |  |  | 26658 |
| Weighted mean |  |  | 8.5 | 8.5 | 14.1 | 13.0 |  | 114202 |
| 2007 | 1 |  | 4.0 | 9.0 | 12.0 |  |  | 829 |
|  | 2 |  | 4.0 | 9.0 | 12.0 |  |  | 244 |
|  | 3 |  | 12.0 | 17.0 | 13.0 | 17.0 |  | 16603 |
|  | 4 | 5.1 | 10.9 | 13.5 | 16.3 |  |  | 66150 |
| Weighted mean |  | 5.1 | 11.1 | 13.8 | 15.5 | 17.0 |  | 83826 |
| 2008 | 1 |  | 4.2 | 7.8 | 10.3 | 10.0 |  | 2930 |
|  | 2 |  | 3.9 | 7.5 | 8.7 |  |  | 52 |
|  | 3 | 2.0 | 11.1 | 11.4 | 12.9 | 14.6 |  | 21759 |
|  | 4 | 3.7 | 10.4 | 13.1 | 13.8 | 14.0 |  | 36362 |
| Weighted mean |  | 3.7 | 9.6 | 11.9 | 13.2 | 14.3 |  | 61102 |

Table 8.2.4. North Sea Sprat. Sampling for biological parameters in 2008.

| Country | Quarter | Landings ('000 tonnes) | No. samples | No. measured | $\begin{array}{r} \text { No. } \\ \text { aged } \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 1 | 1.616 | 3 | 457 | 95 |
|  | 2 | 0.051 | 5 | 47 | 0 |
|  | 3 | 21.787 | 9 | 957 | 428 |
|  | 4 | 36.129 | 6 | 689 | 239 |
|  | Total | 59.583 | 23 | 2150 | 762 |
| UK (England \& Wales) | 1 | 0.032 |  |  |  |
|  | 2 | * |  |  |  |
|  | 3 |  |  |  |  |
|  | 4 | 0.008 |  |  |  |
|  | Total | 0.040 |  |  |  |
| UK (Scotland) | 1 |  |  |  |  |
|  | 2 |  |  |  |  |
|  | 3 |  |  |  |  |
|  | 4 | 0.192 | 1 | 244 | 38 |
|  | Total | 0.192 | 1 | 244 | 38 |
| Norway | 1 | 1.266 | 3 | 300 | 198 |
|  | 2 |  |  |  |  |
|  | 3 |  |  |  |  |
|  | 4 |  |  |  |  |
|  | Total | 1.266 | 3 | 300 | 198 |
| Total North Sea |  | 60.889 | 26 | 2450 | 960 |

Table 8.3.1. North Sea sprat. Abundance indices by age from IBTS (February) from 1984-2009

| Year | Age |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 2 | 3 | 4 | $5+$ | Total |
| 1984 | 233.76 | 329.00 | 39.61 | 6.20 | 0.29 | 608.86 |
| 1985 | 376.10 | 195.48 | 26.76 | 3.80 | 0.35 | 602.49 |
| 1986 | 44.19 | 73.54 | 22.01 | 1.23 | 0.24 | 141.21 |
| 1987 | 542.24 | 66.28 | 19.14 | 1.92 | 0.24 | 629.82 |
| 1988 | 98.61 | 884.07 | 61.80 | 6.99 | 0.00 | 1051.46 |
| 1989 | 2314.22 | 476.29 | 271.85 | 5.47 | 1.65 | 3069.48 |
| 1990 | 234.94 | 451.98 | 102.16 | 28.06 | 2.22 | 819.37 |
| 1991 | 676.78 | 93.38 | 23.33 | 2.63 | 0.12 | 796.24 |
| 1992 | 1060.78 | 297.69 | 43.25 | 7.23 | 0.53 | 1409.48 |
| 1993 | 1066.83 | 568.53 | 118.42 | 6.07 | 0.34 | 1760.19 |
| 1994 | 2428.36 | 938.16 | 92.16 | 3.59 | 0.50 | 3462.77 |
| 1995 | 1224.89 | 1036.40 | 87.33 | 2.52 | 0.76 | 2351.90 |
| 1996 | 186.13 | 383.53 | 146.84 | 18.28 | 0.74 | 735.53 |
| 1997 | 591.86 | 411.95 | 179.55 | 15.52 | 2.24 | 1201.13 |
| 1998 | 1171.05 | 1456.51 | 305.91 | 15.75 | 3.38 | 2952.60 |
| 1999 | 2534.53 | 562.10 | 80.35 | 4.83 | 0.45 | 3182.25 |
| 2000 | 1058.20 | 851.58 | 274.71 | 43.89 | 0.88 | 2229.27 |
| 2001 | 883.06 | 1057.00 | 185.47 | 17.55 | 0.35 | 2143.42 |
| 2002 | 1 | 152.33 | 812.45 | 91.63 | 11.93 | 0.38 |
| 2003 | 1842.26 | 309.92 | 44.49 | 2.21 | 0.04 | 2068.72 |
| 2004 | 1593.89 | 495.70 | 78.24 | 3.50 | 1.54 | 2198.92 |
| 2005 | 3053.46 | 267.89 | 36.39 | 0.87 | 0.00 | 3358.60 |
| 2006 | 421.80 | 1212.87 | 92.38 | 8.26 | 0.07 | 1735.39 |
| 2007 | 1053.68 | 1339.83 | 274.81 | 11.18 | 0.01 | 2679.52 |
| 2008 | 1432.45 | 769.17 | 96.89 | 6.86 | 0.02 | 2305.38 |
| $2009 *$ | 3468.18 | 251.18 | 23.60 | 1.71 | 0.46 | 3745.13 |

[^7]Sprat catches 2008, 1st Quarter


Figure 8.1.1a Sprat catches in the North Sea (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available). a.: 1st quarter

Sprat catches 2008, 2nd Quarter


Figure 8.1.1b Sprat catches in the North Sea (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available). b.: 2nd quarter

## Sprat catches 2008, 3rd Quarter



Figure 8.1.1c Sprat catches in the North Sea (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available). c.: 3rd quarter

## Sprat catches 2008, 4th Quarter



Figure 8.1.1d Sprat catches in the North Sea (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available). d.: 4th quarter

## Sprat catches 2008, All Quarters



Figure 8.1.2
Sprat catches in the North Sea (in tonnes) in 2008 by statistical rectangle. Working group estimates (if available). e: all quarters

## Sprat 1-ringers IBTS 1st Quarter 2009



Figure 8.3.1a
Distribution of 1-ringers in the IBTS (February) 2009 in the North Sea and Division IIIa (Mean number per hour per rectangle).

## Sprat 2-ringers IBTS 1st Quarter 2009



Figure 8.3.1b
Distribution of 2-ringers in the IBTS (February) 2009 in the North Sea and Division IIIa (Mean number per hour per rectangle).

## Sprat 3+ ringers IBTS 1st Quarter 2009



Figure 8.3.1c
Distribution of 3-ringers in the IBTS (February) 2009 in the North Sea and Division IIIa (Mean number per hour per rectangle).


Figure 8.3.2 North Sea Sprat. Abundance (upper figure, in millions) and biomass (lower figure, in 1000 t) per statistical rectangle as obtained by the herring acoustic survey 2008.


Figure 8.3.3 North Sea sprat. Normalized IBTS Q1, Quarter 3 and HERAS acoustic survey indices for the abundance of North Sea sprat. All indices are normalized by their geometric mean over the common period from 2004-2008. For the IBTS survey in Q1-2009 the standard deviation is presented with error bars.


Figure 8.3.4. North Sea sprat. Normalized IBTS Q1, Quarter 3 and HERAS acoustic survey indices for age 1 North Sea sprat. All indices are normalized by their geometric mean over the common period from 2004-2008.


Figure 8.3.5 North Sea sprat. Normalized cumulative-distribution of the per-haul contribution to the IBTS q1 sprat age 1 survey index. The 300-450 individual hauls contributions to the IBTS index in each year are sorted by size and then aggregated to calculate a cumulativedistribution. The plot shows only the contribution for the 30 largest hauls. Numbers on each line indicate the year for the survey.


Figure 8.3.6
North Sea sprat. Time series of IBTS Q1 index values with estimates of uncertainty. Confidence intervals are estimated by the DATRAS database, based on bootstrapping of the raw haul data.

CPUE Sprat 2007 Q1


Figure 8.3.7a North Sea sprat. Comparison of the sprat CPUE in 2007 Q1 in the North Sea, Kattegat/Skagerrak and the English Channel. The red line encircles the North Sea sprat management stock. Data from the Channel has been sampled by the French IBTS survey.

CPUE Sprat 2008 Q1


Figure 8.3.7b North Sea sprat. Comparison of the sprat CPUE in 2008 Q1 in the North Sea, Kattegat/Skagerrak and the English Channel. The red line encircles the North Sea sprat management stock. Data from the Channel has been sampled by the Dutch IBTS survey.

CPUE Sprat 2009 Q1


Figure 8.3.7c North Sea sprat. Comparison of the sprat CPUE in 2009 Q1 in the North Sea, Kattegat/Skagerrak and the English Channel. The red line encircles the North Sea sprat management stock. Data from the Channel has been sampled by the Dutch IBTS survey.

## 9 Sprat in Division IIIa

### 9.1 The Fishery

### 9.1.1 ICES advice applicable for 2008 and 2009

The ACFM advice on sprat management is that exploitation of sprat will be limited by the restrictions imposed on fisheries for juvenile herring. This is a result of sprat being fished mainly together with juvenile herring. The sprat fishery is controlled by a herring by-catch quota as well as by-catch percentage limits (Norway and Denmark: respectively max $10 \%$ and $20 \%$ by-catch of herring in weight). No advice on sprat TAC has been given in recent years.

For 2008 the sprat TAC was set at 52000 t . The by-catch of herring for the EU fleet was 11470 t . For 2009, the TAC for sprat is set at 52000 t and the by-catch quota of herring for the EU fleet at 8373 t .

### 9.1.2 Landings

The total landings decreased from 15700 t in 2007 to 9100 t in 2008 (Table 9.1.1) which is the lowest landings reported. The table presents the landings from 1996 onwards. The data from 1996 and onwards are considered reliable in this context due to the implementation of the new Danish monitoring scheme. The data prior to 1996 can be found in the HAWG report from 2006 (ICES 2006/ACFM:20).

In general, there were sprat landings in all quarters (Table 9.1.2, see Figures 8.1.18.1.2). In 2008 more than $80 \%$ of the total landings were taken in the 1st and 4th quarter. In the Norwegian fishery landings were taken in the 1st and 4th quarter, all as part of the fishery for "anchovy"-production (large sprat).

### 9.1.3 Fleets

Fleets from Denmark, Norway and Sweden carry out the sprat fishery in Division IIIa.

The Danish sprat fishery consists of trawlers using a 16 mm mesh size codend, and all landings are used for fishmeal and oil production. Some of the sprat landings from Denmark and Sweden are by-catches from the herring fishery using 32 mm mesh size codends. There is a Swedish fishery (mainly pelagic trawlers, but also a few purse seiners) directed at herring for human consumption, with by-catches of sprat.

The Norwegian sprat fishery in Division IIIa is a coastal/fjord purse seine fishery for human consumption.

### 9.1.4 Regulations and their effects

Sprat cannot be fished without by-catches of herring except in years with high sprat abundance or low herring recruitment. Management of this stock should consider management advice given for herring in Subarea IV, Division VIId, and Division IIIa.

Most sprat catches are taken in an industrial fishery where catches are limited by herring by-catch restrictions.

### 9.1.5 Changes in fishing technology and fishing patterns

No changes in fishing technology and fishing patterns of importance for the sprat fisheries in IIIa have been reported.

The new Danish quota system (IOK) has also been implemented in 2008 in IIIa (c.f. section 8.1.3).

### 9.2 Biological Composition of the Catch

### 9.2.1 Catches in number and weight-at-age

In 2008 the total numbers of sprat is at the same level as in 2006 and 2007 (Table 9.2.1). In 2008 the majority of the landing (in numbers) is 0 -year olds, contribution $44 \%$ of the total number of landed sprat. This is the highest 0 -year contribution in the period from 2000 and onwards. 1- and 2-year olds contribute respectively $23 \%$ and $19 \%$ to the landed numbers of sprat. Landings of 5+ age group was in 2008 (2\%) the highest in the period given. Data for 1996-2003 is presented in ICES CM 2007/ACFM:11.

Denmark and Sweden provided biological samples from all quarters. No Norwegian samples were collected. Landings in 2008, for which samples were collected, were raised using a combination of Swedish and Danish samples, without any differentiation in types of fleets. Details on the sampling for biological data per country, area and quarter are shown in Table 9.2.3. Mean weight-at-age ( g ) in the catches are presented by quarter in Table 9.2.2. Mean-weight-at-age for all ages is in the same order as the previous years, except for 2007 where the mean weight-at-age were the largest in the period. Mean weights-at-age for 1996-2003 are presented in ICES CM 2005/ACFM:16.

### 9.3 Fishery-independent information

Acoustic estimates of sprat have been available from the ICES co-ordinated Herring Acoustic Surveys in Division IIIa since 1996. At the time of the surveys, sprat has mainly been recorded in the Kattegat (ICES CM 2009/LRC:02).

In 2008 sprat was only observed in the Kattegat (ICES squares 41G1-G2, 42G0-G2, 43G0-G1 and 44G1). The total abundance was estimated to 775 million individuals - a decrease from 6319 million sprat in 2007. The biomass was estimated to 12000 t . Two-year old sprat dominated ( $\sim 60 \%$ ) and half of them were immature (ICES CM 2009/LRC:02).

The IBTS (February) sprat indices for 1984-2009 are presented in Table 9.3.1.
The preliminary total IBTS index for 2009 doubled compared to the 2008 index mainly due to high abundance of 1-group sprat.

### 9.4 Mean weight-at-age and length-at-maturity

Data on maturity by age, mean weight- and length-at-age during the 2008 summer acoustic survey are presented in the PGIPS report (ICES 2009/LRC:02).

### 9.5 Recruitment

For this stock, the IBTS index for 1-group sprat in the first quarter is considered the most suitable recruitment index (Table 9.3.1). The 1-group index for 2009 is well above the average for the time series, and makes $86 \%$ of the total index. This is the
highest dominance of 1-group ever observed. In 2008 the 1-group index contributed less than $10 \%$ of the total index. The procedure for the survey did not differ from previous years. However, the index does not fully reflect strong and weak cohorts seen in the catch. This has also been expressed in a previous working group report (ICES 1998/ACFM:14), and may be linked to difficulties in age determination and/or methodological issues related to the way the indices are estimated (see 3.1.7).

### 9.6 State of the Stock

No assessment of the sprat stock in Division IIIa has been presented since mid1980ies and this year is no exception. Various methods have been explored without success (ICES CM 2007/ACFM:11).

The signal in the IBTS (February) index for 2009 indicates an increase in the sprat stock due to a good incoming year class.

### 9.7 Projection of Catch and Stock

No assessment is presented for this stock.

### 9.8 Reference Points

No precautionary reference points are defined for this stock.

### 9.9 Management Considerations

Sprat in Division IIIa is a short-lived species with large inter-annual fluctuations in stock biomass. The natural inter-annual variability in stock abundance, mainly driven by recruitment variability, is high and does not appear to be strongly influenced by the observed levels of fishing effort.

The sprat has mainly been fished together with herring. The human consumption fishery only takes a minor proportion of the total catch. Within the current management regime, where there is a by-catch ceiling limitation of herring as well as bycatch percentage limits, the sprat fishery is controlled by these factors. In the last years the sprat fisheries has not been limited by the sprat quota, since this quota has not been taken.

### 9.10 Ecosystem Considerations

No information of the ecosystem and the accompanying considerations are known at present. In the adjacent North Sea Multispecies investigations have demonstrated that sprat is one of the important prey species in the North Sea ecosystem, as a prey species for both fish and seabirds. Many of the plankton feeding fish have recruited poorly in recent years (e.g. herring, sandeel, Norway pout). The implications for sprat in IIIa are at present unknown.

### 9.11 Changes in the environment

Temperatures in the area have increased over the last years. It is considered that this may have implications for sprat.

Table 9.1.1 Division Illa sprat. Landings in ('000 t) 1996-2008.
(Data provided by Working Group members). These figures do not in all cases correspond to the official statistics and cannot be used for management purposes.

| Year | Skagerrak |  |  |  | Kattegat |  |  | Div. Illa total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Denmark | Sweden | Norway | Total | Denmark | Sweden | Total |  |
| 1996 | 7.0 | 3.5 | 1.0 | 11.5 | 3.4 | 3.1 | 6.5 | 18.0 |
| 1997 | 7.0 | 3.1 | 0.4 | 10.5 | 4.6 | 0.7 | 5.3 | 15.8 |
| 1998 | 3.9 | 5.2 | 1.0 | 10.1 | 7.3 | 1.0 | 8.3 | 18.4 |
| 1999 | 6.8 | 6.4 | 0.2 | 13.4 | 10.4 | 2.9 | 13.3 | 26.7 |
| 2000 | 5.1 | 4.3 | 0.9 | 10.3 | 7.7 | 2.1 | 9.8 | 20.1 |
| 2001 | 5.2 | 4.5 | 1.4 | 11.2 | 14.9 | 3.0 | 18.0 | 29.1 |
| 2002 | 3.5 | 2.8 | * | 6.3 | 9.9 | 1.4 | 11.4 | 17.7 |
| 2003 | 2.3 | 2.4 | 0.8 | 5.6 | 7.9 | 3.1 | 10.9 | 16.5 |
| 2004 | 6.2 | 4.5 | 1.1 | 11.8 | 8.2 | 2.0 | 10.2 | 22.0 |
| 2005 | 12.1 | 5.7 | 0.7 | 18.5 | 19.8 | 2.1 | 21.8 | 40.3 |
| 2006 | 1.2 | 2.8 | 0.3 | 4.3 | 6.6 | 1.6 | 8.2 | 12.5 |
| 2007 | 1.4 | 2.8 | 1.6 | 5.9 | 8.5 | 1.3 | 9.8 | 15.7 |
| 2008 | 0.3 | 1.5 | 0.9 | 2.6 | 5.6 | 0.9 | 6.5 | 9.1 |

* < 50 t

Table 9.1.2. Division Illa sprat. Landings of sprat (' O 00 t ) by quarter
by countries, 2000-2008. Data for 1996-1999 in ICES CM 2007/ACFM:11 (Data provided by the Working Group members)

|  | Quarter | Denmark | Norway | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 1 | 4.1 | 0.1 | 2.3 | 6.5 |
|  | 2 |  |  | 1.9 | 1.9 |
|  | 3 | 4.8 | 0.1 |  | 4.9 |
|  | 4 | 3.8 | 0.7 | 2.3 | 6.8 |
|  | Total | 12.7 | 0.9 | 6.4 | 20.0 |
| 2001 | 1 | 2.5 |  | 2.6 | 5.2 |
|  | 2 | 6.6 |  | 0.1 | 6.7 |
|  | 3 | 10.2 |  | 0.1 | 10.2 |
|  | 4 | 0.9 | 1.4 | 4.8 | 7.1 |
|  | Total | 20.2 | 1.4 | 7.6 | 29.1 |
| 2002 | 1 | 3.8 |  | 1.4 | 5.2 |
|  | 2 | 2.1 |  | 0.4 | 2.4 |
|  | 3 | 5.9 |  | 0.1 | 6.0 |
|  | 4 | 1.7 |  | 2.4 | 4.1 |
|  | Total | 13.4 |  | 4.3 | 17.7 |
| 2003 | 1 | 3.5 | 0.1 | 1.7 | 5.3 |
|  | 2 | 0.6 |  | 0.8 | 1.4 |
|  | 3 | 1.0 |  | 0.7 | 1.7 |
|  | 4 | 5.0 | 0.8 | 2.3 | 8.1 |
|  | Total | 10.2 | 0.8 | 5.5 | 16.5 |
| 2004 | 1 | 3.1 |  | 1.4 | 4.5 |
|  | 2 | 0.6 |  | 0.9 | 1.5 |
|  | 3 | 3.7 |  | 0.4 | 4.1 |
|  | 4 | 6.9 | 1.1 | 3.8 | 11.9 |
|  | Total | 14.4 | 1.1 | 6.5 | 22.0 |
| 2005 | 1 | 6.5 |  | 1.7 | 8.1 |
|  | 2 | 4.6 |  | 0.1 | 4.7 |
|  | 3 | 18.6 | 0.7 | 0.8 | 20.1 |
|  | 4 | 2.1 |  | 5.2 | 7.3 |
|  | Total | 31.9 | 0.7 | 7.7 | 40.3 |
| 2006 | 1 | 5.4 | 0.2 | 2.7 | 8.3 |
|  | 2 | 0.2 |  | 0.2 | 0.3 |
|  | 3 | 1.3 |  | 0.1 | 1.4 |
|  | 4 | 0.9 | 0.1 | 1.5 | 2.5 |
|  | Total | 7.8 | 0.3 | 4.4 | 12.5 |
| 2007 | 1 | 2.3 | 0.4 | 0.4 | 3.1 |
|  | 2 | 0.7 |  | 0.6 | 1.3 |
|  | 3 | 5.1 | * | 0.2 | 5.4 |
|  | 4 | 1.8 | 1.2 | 3.0 | 5.9 |
|  | Total | 9.9 | 1.6 | 4.2 | 15.7 |
| 2008 | 1 | 2.3 | 0.2 | 0.6 | 3.1 |
|  | 2 | 0.7 |  | 0.4 | 1.0 |
|  | 3 | 0.4 |  | 0.2 | 0.6 |
|  | 4 | 2.5 | 0.7 | 1.2 | 4.4 |
|  | Total | 5.8 | 0.9 | 2.4 | 9.1 |



Table 9.2.2. Division Illa sprat. Quarterly mean weight-at-age (g) in the landings for the years 2000-2007. The equivalent data for 1996-2003 can be found in ICES CM 2007 /ACFM: 11. (Danish and Swedish data)

| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Quarter | 0 | 1 | 2 | 3 | 4 | $5+$ |
| 2004 | 1 |  | 4.6 | 14.6 | 17.8 | 17.3 | 17.3 |
|  | 2 |  | 7.0 | 13.6 | 16.7 | 17.0 | 19.5 |
|  | 3 | 3.0 | 14.1 | 16.7 | 20.0 | 21.4 |  |
|  | 4 | 3.5 | 16.8 | 19.9 | 22.2 | 20.9 | 28.0 |
| Weighted mean |  | 3.5 | 10.4 | 16.3 | 18.4 | 17.8 | 17.9 |
| 2005 | 1 |  | 3.0 | 14.6 | 16.3 | 20.3 | 21.1 |
|  | 2 |  | 5.4 | 11.7 | 26.8 |  |  |
|  | 3 | 2.9 | 11.9 | 14.6 | 15.4 | 11.0 |  |
|  | 4 | 3.3 | 13.1 | 19.1 | 20.1 | 21.1 | 23.1 |
| Weighted mean |  | 5.0 | 7.6 | 15.4 | 17.1 | 17.2 | 21.5 |
| 2006 | 1 |  | 5.0 | 12.2 | 15.4 | 15.2 | 18.5 |
|  | 2 |  | 7.0 | 13.3 | 16.3 | 22.0 |  |
|  | 3 |  | 11.2 | 17.4 | 20.3 | 18.6 | 22.8 |
|  | 4 | 4.3 | 16.1 | 19.6 | 21.4 | 23.8 | 26.6 |
| Weighted mean |  | 4.3 | 6.8 | 13.6 | 16.8 | 16.1 | 19.4 |
| 2007 | 1 |  | 2.3 | 12.3 | 16.3 | 17.0 | 25.2 |
|  | 2 |  | 6.1 | 17.1 | 20.6 | 21.9 | 20.4 |
|  | 3 |  | 12.0 | 13.0 | 17.0 | 17.6 |  |
|  | 4 | 7.9 | 14.1 | 20.3 | 23.4 | 22.6 | 26.2 |
| Weighted mean |  | 7.9 | 11.5 | 15.9 | 18.4 | 19.3 | 25.2 |
| 2008 | 1 |  | 5.6 | 11.7 | 15.5 | 18.1 | 18.3 |
|  | 2 |  | 8.0 | 12.5 | 17.1 | 19.3 | 22.2 |
|  | 3 | 3.4 | 7.9 | 21.1 | 21.5 | 25.3 | 22.5 |
|  | 4 | 3.4 | 9.2 | 20.7 | 21.4 | 25.2 | 22.8 |
| Weighted mean |  | 3.4 | 9.0 | 13.3 | 16.9 | 19.5 | 20.0 |

Table 9.2.3 Division Illa sprat. Sampling commercial landings for biological samples in 2008.

| Country | Quarter | Landings (tonnes) | No. <br> samples | $\begin{array}{r} \text { No. } \\ \text { meas. } \end{array}$ | No. aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 1 | 2253 | 4 | 529 | 153 |
|  | 2 | 671 | 5 | 163 | 56 |
|  | 3 | 446 | 1 | 77 |  |
|  | 4 | 2459 | 3 | 336 | 147 |
|  | Total | 5829 | 13 | 1105 | 356 |
| Norway | 1 | 190 |  |  |  |
|  | 2 |  |  |  |  |
|  | 3 |  |  |  |  |
|  | 4 | 704 |  |  |  |
|  | Total | 894 |  |  |  |
| Sweden | 1 | 637 | 3 | 697 | 697 |
|  | 2 | 350 |  |  |  |
|  | 3 | 189 |  |  |  |
|  | 4 | 1213 | 11 | 539 | 539 |
|  | Total | 2389 | 14 | 1236 | 1236 |
| Denmark |  | 5829 | 51 | 4501 | 955 |
| Norway |  | 894 |  |  |  |
| Sweden |  | 2389 | 18 | 1226 | 1225 |
|  | Total | 9112 | 69 | 5727 | 2180 |

Table 9.3.1. Division Illa sprat. IBTS (February) indices of sprat per age group 1984-2009.

| Year | No Rect | No hauls | Age Group |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 1 | 2 | 3 | 4 | 5+ | Total |
| 1984 | 15 | 38 | 5675.45 | 868.88 | 205.10 | 79.08 | 63.57 | 6892.08 |
| 1985 | 14 | 38 | 2157.76 | 2347.02 | 392.78 | 139.74 | 51.24 | 5088.54 |
| 1986 | 15 | 38 | 628.64 | 1979.24 | 2034.98 | 144.19 | 37.53 | 4824.58 |
| 1987 | 16 | 38 | 2735.92 | 2845.93 | 3003.22 | 2582.24 | 156.64 | 11323.95 |
| 1988 | 13 | 38 | 914.47 | 5262.55 | 1485.07 | 2088.05 | 453.13 | 10203.26 |
| 1989 | 14 | 38 | 413.94 | 911.28 | 988.95 | 554.53 | 135.79 | 3004.48 |
| 1990 | 15 | 38 | 481.02 | 223.89 | 64.93 | 61.11 | 45.69 | 876.65 |
| 1991 | 14 | 38 | 492.50 | 726.82 | 698.11 | 128.36 | 375.44 | 2421.23 |
| 1992 | 16 | 38 | 5993.64 | 598.71 | 263.97 | 202.90 | 76.04 | 7135.25 |
| 1993 | 16 | 38 | 1589.92 | 4168.61 | 907.43 | 199.32 | 239.64 | 7104.92 |
| 1994 | 16 | 38 | 1788.86 | 715.84 | 1050.87 | 312.65 | 70.11 | 3938.32 |
| 1995 | 17 | 38 | 2204.07 | 1769.53 | 35.19 | 44.96 | 4.23 | 4057.98 |
| 1996 | 15 | 38 | 199.30 | 5515.42 | 692.78 | 111.98 | 173.75 | 6693.23 |
| 1997 | 16 | 41 | 232.65 | 391.23 | 1239.13 | 139.14 | 134.51 | 2136.67 |
| 1998 | 15 | 39 | 72.25 | 1585.22 | 619.76 | 1617.71 | 521.52 | 4416.46 |
| 1999 | 16 | 42 | 4534.96 | 355.24 | 249.86 | 44.25 | 313.52 | 5497.83 |
| 2000 | 16 | 41 | 292.32 | 737.80 | 59.69 | 51.79 | 23.21 | 1164.80 |
| 2001 | 16 | 42 | 6539.48 | 1144.34 | 676.71 | 92.37 | 45.87 | 8498.77 |
| 2002 | 16 | 42 | 1180.52 | 1035.71 | 89.96 | 58.85 | 12.93 | 2377.96 |
| 2003 | 17 | 46 | 462.64 | 1247.49 | 1172.13 | 382.29 | 123.17 | 3387.72 |
| 2004 | 16 | 41 | 402.87 | 49.00 | 156.62 | 86.57 | 27.48 | 722.54 |
| 2005 | 17 | 50 | 3314.17 | 1563.16 | 470.84 | 837.09 | 538.37 | 6723.63 |
| 2006 | 17 | 45 | 1323.59 | 11855.76 | 1753.92 | 299.05 | 159.23 | 15391.55 |
| 2007 | 18 | 46 | 774.11 | 306.63 | 250.81 | 42.08 | 13.74 | 1387.37 |
| 2008 | 17 | 46 | 150.85 | 982.68 | 132.54 | 228.48 | 107.70 | 1602.26 |
| 2009* | 17 | 46 | 2686.72 | 124.46 | 259.15 | 29.60 | 37.43 | 3137.36 |

* Preliminary


## 10 Stocks with insufficient data

Two stocks with very low research intensity were poorly described in previous reports in devoted sections or chapters. These were Clyde herring (Section 5.11 in ICES 2005a) and sprat in VIId,e (Section 9 in ICES 2005). The advice on these stocks cannot be improved at present. In this section only the times series are maintained. For most recent advice refer to the appropriate sections in the HAWG report (ICES CM 2005/ACFM:18).

There was sampling of the catch in 2008 for Clyde herring, with one sample available in quarter 4. The catch of Clyde herring in 2008 was low (Table 10.1). The 2008 Clyde herring catch was on a par with the 2007 catch and, again, slightly higher than in recent years.

The catches of sprat in VIId and VIIe were nearly doubled in 2008 compared to the past years (Table 10.2). Landings have not been at the level of 2008 since 1999.

Table 10.1 Herring from the Firth of Clyde. Catch in tonnes by country, 1955-2008. Spring and autumn-spawners combined.

| Year | 1955 | 1956 | 1957 | 1958 | 1959 | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| All Catches |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 4050 | 4848 | 5915 | 4926 | 10530 | 15680 | 10848 | 3989 | 7073 | 14509 | 15096 | 9807 | 7929 | 9433 |


| Year | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| All Catches |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Total | 10594 | 7763 | 4088 | 4226 | 4715 | 4061 | 3664 | 4139 | 4847 | 3862 | 1951 |


| Year | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Scotland | 2506 | 2530 | 2991 | 3001 | 3395 | 2895 | 1568 | 2135 | 2184 | 713 | 929 | 852 | 608 | 392 |
| Other UK | - | 273 | 247 | 22 | - | - | - | - | - | - | - | 1 | - | 194 |
| Unallocated ${ }^{1}$ | 262 | 293 | 224 | 433 | 576 | 278 | 110 | 208 | 75 | 18 | - | - | - | - |
| Discards | 1253 | 1265 | $2308{ }^{3}$ | $1344{ }^{3}$ | 6793 | 4394 | $245{ }^{4}$ | -2 | -2 | -2 | -2 | -2 | -2 | -2 |
| Agreed TAC |  |  | 3000 | 3000 | 3100 | 3500 | 3200 | 3200 | 2600 | 2900 | 2300 | 1000 | 1000 | 1000 |
| Total | 4021 | 4361 | 5770 | 4800 | 4650 | 3612 | 1923 | 2343 | 2259 | 731 | 929 | 853 | 608 | 586 |


| Year | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Scotland | 598 | 371 | 779 | 16 | 1 | 78 | 46 | 88 | - | - | + | 163 | 54 |
| Other UK | 127 | 475 | 310 | 240 | 0 | 392 | 335 | 240 | - | 318 | 512 | 458 | 622 |
| Unallocated ${ }^{1}$ | - | - | - | - | - | - | - | - | - | - | - | - |  |
| Discards | - | - | - | - | - | - | - | - | - | - | - | - |  |
| Agreed TAC | 1000 | 1000 | 1000 | 1000 | 1000 | 1000 | 1000 | 1000 | 1000 | 1000 | 1000 | 800 | 800 |
| Total | 725 | 846 | 1089 | 256 | 1 | 480 | 381 | 328 | 0 | 318 | 512 | 621 | 676 |

${ }^{1}$ Calculated from estimates of weight per box and in some years estimated by-catch in the sprat fishery
${ }^{3}$ Based on sampling.
${ }^{2}$ Reported to be at a low level, assumed to be zero, for 1989-1995.
${ }^{4}$ Estimated assuming the same discarding rate as in 1986

Table 10.2. Sprat VIId,e. Nominal catches in tonnes of sprat in VIId,e from 1985-2008

|  | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  | 15 | 250 | 2529 | 2092 | 608 |  |  |
| France | 14 |  | 23 | 2 | 10 |  |  | 35 |
| Netherlands |  |  |  |  |  |  |  |  |
| UK (Engl.\&Wales) | 3771 | 1163 | 2441 | 2944 | 1319 | 1508 | 2567 | 1790 |
| Total | 3785 | 1178 | 2714 | 5475 | 3421 | 2116 | 2567 | 1825 |
| Country | 1993 | 1994 | 1995 | 1996 | 1997 | 1998* | 1999* | 2000* |
| Denmark |  |  |  |  |  |  |  |  |
| France | 2 | 1 | 0 |  |  |  |  | 18 |
| Netherlands |  |  |  |  |  |  | 1 | 1 |
| UK (Engl.\&Wales) | 1798 | 3177 | 1515 | 1789 | 1621 | 2024 | 3559 | 1692 |
| Total | 1800 | 3178 | 1515 | 1789 | 1621 | 2024 | 3560 | 1711 |
| Country | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| Denmark |  |  |  |  |  |  |  |  |
| France |  |  |  |  |  |  |  |  |
| Netherlands |  |  |  |  |  |  |  |  |
| UK (Engl.\&Wales) | 1349 | 1196 | 1377 | 836 | 1635 | 1974 | 1819 | 3366 |
| Total | 1349 | 1196 | 1377 | 836 | 1635 | 1974 | 1819 | 3366 |
| * Preliminary |  |  |  |  |  |  |  |  |

## 11 Working Documents

WD 01 Beggs,S., Schon,P.J., McCurdy,W., Peel,J., McCorriston, P., McCausland,I.: Seasonal Origin of 1-ring+ Herring in the Irish Sea (VIIaN) Management Area During the Annual Acoustic Survey. AFBI, Belfast, N. Ireland, UK.

WD 02 Brunel, Dickey-Collas, M. :Recruitment-environment relationships in North Sea herring revisited : looking for limiting factors using quantile regression. Wageningen IMARES, IJmuiden, The Netherlands.

WD 03 Aloysius, T.M., van Helmond Harriët M.J. van Overzee: Estimates of discarded herring by Dutch flagged vessels 2003-2008 Wageningen IMARES, IJmuiden, The Netherlands.

WD 04 Dickey-Collas, M., Bolle, L.J., van Beek J.K.L., Erftemeijer P.L.A.: How variable is the interannual transport of herring larvae in the southern North Sea?, Wageningen IMARES, IJmuiden, The Netherlands.

WD 05 Van Damme, C.J.G., Dickey-Collas, M., Rijnsdorp,A.D., Kjesbu, O.S.: Fecundity, atresia and spawning strategies of Atlantic herring. Wageningen IMARES, IJmuiden, The Netherlands.

WD 06 Gröhsler, T.: Fisheries \& Stock assessment data in the Western Baltic in 2008. vTI-OSF, Rostock, Germany.

WD 07 Skagen, D.: The assessment of North Sea sprat: Is length structured models a way forward? Analysis of data and runs with the program LCS. Institute of Marine Research Bergen, Norway February 2009

WD 08 Harma, C., Clarke M., Brophy D.: Relative strength of autumn spawners in Celtic Sea herring over time, Marine Institute, Rinville, Oranmore, Galway, Ireland.
WD 09 Clarke, M., Egan A, Molloy J.: Developing a recruit index for Celtic Sea herring, Marine Institute, Rinville, Oranmore, Galway, Ireland.

WD 10 Bierman, S. M., Dickey-Collas,M., van Damme, C. J. G.,. van Overzee, H. M. J., Pennock, I M. G. Tribuhl, S. V. Wageningen IMARES, IJmuiden, The Netherlands. L. A. W. Clausen, L.A.W.: Mixing of North Sea herring spawning components in the summer catch, is there a consistent pattern?

WD 11 Dickey-Collas, M., Clarke, M., Slotte, A., (Conveners): Linking Herring': do we really understand plastic? Wageningen IMARES, IJmuiden, The Netherlands. Marine Institute, Galway, Ireland. A. Slotte: IMR, Bergen, Norway.
WD 12 Rohlf, N., Gröger, J.: Report of the herring larvae surveys in the North Sea in 2008/2009. vTI-SF, Hamburg, Germany.

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## ANNEX 1: List of Participants

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17-25 March 2009
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## Annex 2 - Recommendations

HAWG 2009 makes the following recommendations:

| RECOMMENDATION | ACTION |
| :--- | :--- |
| The FLICA assessment method cannot be maintained <br> in its current form, due to its inclusion of the proprie- <br> tary NAG optimisation libraries. HAWG and ICES <br> should develop a roadmap for the succession of this <br> method, with the aim of employing an alternative im- <br> plementation or assessment model in the next round of |  |
| benchmark assessments. |  |

2. The 1-quarter trawl survey, using GOV trawl, conducted in 2009, should continue in subsequent years.
3. The time allocated to VIIj in the $\mathrm{q}-4$ Celtic Sea acoustic survey has rarely encountered substantial herring abundance. Sacrificing this VIIj acoustic ship time would not jeopardize the existing acoustic index. However the ship time saved could be reallocated to the $\mathrm{q}-1$ trawl survey mentioned in point 2 above.

The WESTHER project has demonstrated that fish of several different stocks mix in the VIaN area. HAWG

HAWG, PGIPS, PGCCDBS recommends the splitting of survey abundances by spawning season be implemented in 2009 and subsequent years. This should be implemented for each of the three constituent surveys of the Malin Shelf.

HAWG recommends that discrepancies at the area lev-
ICES InterCatch el in the output of conventional used systems and InterCatch should be elucidated in detail between stockcoordinators and ICES InterCatch team. Furthermore, routines should be implemented in InterCatch to report on CATON, WECA and CANUM for area IIIa, and for NSAS and WBSS spawners separately.

## Annex 3-Stock Annex North Sea Herring

Quality Handbook ANNEX: hawg-her47d3
Stock specific documentation of standard assessment procedures used by ICES.
Stock: North Sea Autumn Spawning Herring (NSAS)
Working Group: $\quad$ Herring Assessment WG for the Area south of $62^{\circ} \mathrm{N}$
Date: 22 March 2009
Authors: C. Zimmermann, J. Dalskov, M. Dickey-Collas,
H. Mosegaard, P. Munk, J. Nichols, M. Pastoors, N. Rohlf, E.J. Simmonds, D. Skagen, N. Payne, M. Payne

## A. General

## A.1. Stock definition:

Autumn spawning herring distributed in ICES area IV, Division IIIa and VIId. Mixing with other stocks occurs especially in Division IIIa (with Western Baltic Spring Spawning herring). Genetic studies have failed to prove that the stock is not one unit (Mariani et al., 2005; Reiss et al., 2009).

## A.2. Fishery

North Sea Autumn Spawners are exploited by a variety of fleets, ranging from small purse seiners to large freezer trawlers, of different nations (Norway, Denmark, Sweden, Germany, The Netherlands, Belgium, France, UK, Faroe Islands). The majority of the fishery takes place in the Shetland-Orkney area and northern North Sea in the $2^{\text {nd }}$ and $3^{\text {rd }}$ quarter, and in the English Channel (Division VIId) in the $4^{\text {th }}$ quarter. Juveniles are caught in Division IIIa and as by-catch in the industrial fishery in the central North Sea. For management purposes, 4 fleets are currently defined: Fleet A is harvesting herring for human consumption in IV and VIId, but includes herring bycatches in the Norwegian industrial fishery; fleet B is the industrial (small mesh, <32 mm mesh size) fleet of EU nations operating in IV and VIId. North Sea Autumn spawners are also caught in IIIa in fleets $C$ (human consumption) and D (small mesh).

## A.3. Ecosystem aspects:

Herring is the key pelagic species in the North Sea and is thus considered to have major impact as prey and predator to most other fish stocks in that area.
The North Sea is semi-enclosed and situated on the continental shelf of Northwestern Europe and is bounded by England, Scotland, Norway, Sweden, Denmark, Germany, the Netherlands, Belgium and France. It covers an area of $\sim 750000 \mathrm{~km}^{2}$ of which the greater part is shallower than 200 m . It is one of the most diverse coastal regions in the world, with a variety of coastal habitats (fjords, estuaries, deltas, banks, beaches, sandbanks and mudflats, marshes, rocks and islands), and four ecological seasons. It is a highly productive ( $>300 \mathrm{gC} \mathrm{m}^{-2} \mathrm{yr}^{-1}$ ) ecosystem but with primary productivity varying considerably across the sea. The highest values of primary productivity occur in the coastal regions, influenced by terrestrial inputs of nutrients, and in
areas such as the Dogger Bank and tidal fronts. Changes observed in trophic structure are indicative of a trend towards a decreasing resilience of this ecosystem. This trend is partially a response to inter-annual changes in the physical oceanography of the North Atlantic.

Herring are an integral and important part of the pelagic ecosystem in the North Sea. As plankton feeders they form an important part of the food chain up to the higher trophic levels. Both as juveniles and as adults they are an important source of food for some demersal fish and for sea mammals. Over the past century the top predator, man, has exerted the greatest influence on the abundance and distribution of herring in the North Sea. Spawning stock biomass has fluctuated from estimated highs of around 4.5 million tonnes in the late 1940s to lows of less than 100000 tonnes in the late 1970s (Simmonds 2007). The species has demonstrated robustness in relation to recovery from such low levels once fishing mortality is curtailed in spite of recruitment levels being adversely affected (Payne et al., 2009, Nash et al., 2009).

Their spawning and nursery areas, being near the coasts, are particularly sensitive and vulnerable to anthropogenic influences. The most serious of these is the ever increasing pressure for marine sand and gravel extraction and the development of wind farms. This has the potential to seriously damage and to destroy the spawning habitat and disturb spawning shoals and destroy spawn if carried out during the spawning season. It also has the potential to destroy traditional spawning grounds which are currently unused but likely to be recolonised. Similarly, trawling at or close to the bottom in known spawning areas can have the same detrimental effects. It is possible that the disappearance of spawning on the western edge of the Dogger bank could well be attributable to such anthropogenic influences.

In more recent years the oil and gas exploration in the North Sea has represented a potential threat to herring spawning although great care has been taken by the industry to restrict their activities in areas and at times of known herring spawning activity.

By-catch and Discard
By-catch consists of the retained 'incidental' catch of non-target species and discard is a deliberately (or accidentally) abandoned part of the catch returned to the sea as a result of economic, legal, or personal considerations. This section therefore deals with these two elements of the fishery, looking specifically at fishery-related issues. Cetacean, seabird and other threatened, rare and iconic species which may form part of a by-catch are considered separately in the next section. All discarding is illegal for Norwegian vessels and slippage and high grading is now illegal for EU vessels if quota is still available and the fish are above minimum landing size.

Incidental Catch: The incidental catch of non-target species in the North Sea pelagic herring fishery in general is considered to be low (Borges et al., 2008). A study by Pierce et al. (2002) investigated incidental catch from commercial pelagic trawlers over the period January to August 2001. The target species, herring, accounted for $98 \%$ by weight of the overall catch with an overall incidental catch of $2.3 \%$ made up of mackerel, haddock, horse mackerel and whiting. However, onboard sampling over 2002 by Scottish and German observers found substantial discards of herring, taken as by-catch in the mackerel fishery over the 3rd and 4th quarters, after herring quotas had been exhausted. This was not found in a study of the Dutch fleet (Borges et al., 2008) when the herring fishery was found to be relatively "clean".

Discards and slipping: The indications are that large-scale discarding is not widespread in the directed North Sea herring fishery. A number of direct-observer surveys have been conducted on Scottish, Dutch and Norwegian pelagic trawlers, (Napier et al, 1999; 2002; Borges et al., 2008). The overall discard rate was less than $5 \%$ of the landed catch. It is likely that there are different discard rates between the specific fishing types. There is disagreement about the amount of slippage compared to discarding by the differing fleets (slippage- fish released from the nets whilst still in the water but still resulting in the mortality of the majority of pelagic fish, discardingfish dumped back into the sea after having been brought on board). For both pursers and trawlers 'poor' fish quality was a significant cause of discarding. The strength of year classes influences discarding behaviour, particularly of undersized fish. The influence of strong herring year classes was apparent in the composition of discards with smaller, younger fish accounting for a high proportion of the fish discarded in 2001. In the mid 2000s the stronger recruitment of mackerel has probably lead to the increase in discarding of smaller mackerel.

Ecosystem Considerations. The incidental non-target fish catch by directed North Sea herring fisheries appears to be low (ca. 2\%), mainly consisting of mackerel when fishing mixed shoals. Thus it is likely that the impact of incidental fish catches is negligible. The discard of unwanted herring, mostly in the form of high-grading to improve catch quality and grade sizes of fish between 2-4 years of age is low and now illegal in both the EU and Norway. Discarding is thought to be reducing.

Interactions with Rare, Protected or charismatic mega fauna: Interactions between the directed North Sea herring fishery with rare, protected or charismatic mega fauna species are, in general, considered to be low. Species which may interact with the fishery are considered below.
Cetacean by-catch: Since 2000, the Sea Mammal Research Unit (SMRU) of St. Andrew's University in Scotland, under contract to DEFRA, has carried out a number of surveys to estimate the level of by-catch in UK pelagic fisheries. SMRU, in collaboration with the Scottish Pelagic Fishermen's Association, placed observers on board thirteen UK vessels for a total of 190 days at sea, covering 206 trawling operations around the UK. No cetacean by-catch was observed in the herring pelagic fishery in the North Sea. Pierce (2002) also reports that no by-catches of marine mammals were observed over 69 studies hauls and considers that the underlying rate for marine mammals in the pelagic fisheries studies (pelagic trawls in IVa and VIa) is no more than 0.05 (i.e. five events per 100 hauls) and may well be considerably lower than this. Consequently, the cetacean by-catch by the pelagic trawl fishery can be regarded as negligible. This was also confirmed by an UK observer programme ended in 2003 (Northridge, pers. Comm.).

Other than the above, there are no reliable estimates of by-catch for pelagic trawl fisheries, though observations have been made and by-catch rates have been established for several fisheries. Data are now collected routinely through the DRF and have yet to be analysed. Kuklik and Skóra (2003) refer to a single record of a harbour porpoise (Phocoena phocoena) by-caught in a herring trawl in the Baltic. Observations in several other pelagic trawl fisheries were reported by Morizur et al. (1999) and Couperus (1997). All appear to agree that incidental catches of cetaceans in the Dutch pelagic trawl fishery are largely restricted to late-winter/early-spring in an area along the continental slope southwest of Ireland, so outside the North Sea.

Seal by-catch: The by-catch of seals in directed pelagic herring fishery in the North Sea is reported to be "very rare" (Aad Jonker, pers. comm.). Independent verification
also confirms this to be so, with perhaps one animal being caught by the whole North Sea fleet a year (Bram Couperus (IMARES, pers. comm.). Northridge (2003) observed 49 seals taken in 312 pelagic trawl tows throughout UK waters and reports that the fishery in North-western Scotland has the highest observed seal by-catch levels of UK pelagic trawl fisheries, possible amounting to dozens per year. Although not confirmed, it was assumed that the majority were grey seal Halichoerus grypus. This species is mainly distributed around the Orkneys and Outer Hebrides - out of a UK population of 129000 , only around 7000 and 5900 are distributed off the Scottish and English North Sea coasts respectively (SCOS, 2002), and so by-catch rates in the North Sea are likely to be substantially less than off the NW Scottish coast. The eastern Atlantic population of the Grey seal is not considered to be threatened.

Other by-catch: Sharks are occasionally caught by pelagic trawlers in the North Sea, although this is rare with a maximum of two fish per trip (Aad Jonker, pers. comm.). Survival rates are apparently high, sharks are released during or after the cod-end is being emptied. The species are unknown, although blue shark Prionace glauca, which preys primarily upon schooling fishes such as anchovies, sardines and herring, are known to have been caught by pelagic trawls off the SW English coast (Bram Couperus (IMARES), pers. comm.). Gannets (Morus bassanus), which frequently dive at and around nets, were observed by Napier et al. (2002) entangled in the nets but were not present in samples. Actual mortality rates of caught gannets have not been assessed in detail, and some have been observed alive after release from the gear. An extrapolation from observed mortalities corresponds to around 560 gannet deaths per year, although this is based on a relatively low sample frame. Seabird by-catch in the North Sea is considered to be comparatively rare. In the NW Scotland, 1-3 birds may be caught, especially in grounds off St. Kilda (Aad Jonker (former freezer trawler skipper), pers. comm.). IMARES observers in the North Sea only recorded one incident of seabird by-catch over 10 trips (Bram Couperus, pers. comm.).

## B. Data

## B.1. Commercial catch:

Commercial catch is obtained from national laboratories of nations exploiting herring in the North Sea. Since 1999 (catch data 1998), these labs have used a spreadsheet to provide all necessary landing and sampling data, which was developed originally for the Mackerel Working Group (WGMHSA) and further adapted to the special needs of the Herring Assessment Working Group. The current version used for reporting the 2007 catch data was v1.6.4. This method is now run in parallel with INTERCATCH, which is maintained by ICES. INTERCATCH is still in development and thus HAWG uses both. The data in the exchange spreadsheets are allocated samples to catch using the SALLOCL-application (Patterson, 1998). This programme gives the needed standard outputs on sampling status and biological parameters. It also clearly documents any decisions made by the species co-ordinators for filling in missing data and raising the catch information of one nation/quarter/area with information from another data set.

In addition, commercial catch and sampling data were stored and processed using the Intercatch-software for the first time during the WG in 2007. While at that time larger discrepancies up to $5 \%$ between the SALLOCL routines and Intercatch did occur, INTERCATCH performed quite well in 2008. The estimates of CANON, CATON and WECA were highly comparable. However INTERCATCH is still not completely satisfactory in terms of flexibility and outputs. Thus both methods are still being used.

The "wonderful table". The following figure explains were the estimates in the wonderful table are derived from:


Transparency of data handling by the Working Group. The current practice of data handling by the Working Group is that the data received by the co-ordinators is available in a folder called "archive". These high-resolution data are not reproduced in the report. The archived data contains the disaggregated dataset (disfad), the allocations of samples to unsampled catches (alloc), the aggregated dataset (sam.out) and (in some cases) a document describing any problems with the data in that year. Since 2007, the corresponding datasets are also stored in Intercatch, where they are accessible to the stock coordinators only.

Current methods of compiling fisheries assessment data. The stock co-ordinator is responsible for compiling the national data to produce the input data for the assessments. In addition to checking the major task involved is to allocate samples of catch numbers, mean length and mean weight-at-age to unsampled catches. There are at present no defined criteria on how this should be done, but the following general process is implemented by the species co-ordinators. Searches are made for appropriate samples by gear (fleet), area and quarter. If an exact match is not available the search will move to a neighbouring area if the fishery extends to this area in the same quarter. More than one sample may be allocated to an unsampled catch, in this case a straight mean or weighted mean of the observations may be used. If there are no samples available the search will move to the closest non-adjacent area by gear (fleet) and quarter, but not in all cases.

The Working Group acknowledges the effort some members have made to provide "corrected" data, which in some cases differ significantly from the officially reported catches. Most of this valuable information is gathered on the basis of personal knowledge of the fishery and good relations between the scientist responsible and the fishermen. In addition the Working Group recognises and would like to highlight the inherent conflict of interest in obtaining details of unallocated catches by country and increasing the transparency of data handling by the Working Group.

## B.2. Biological

Catch-at-age data (catch numbers-at-age, mean weights-at-age in the catch, mean length-at-age) is derived from the raised national figures received from the national laboratories. The data are obtained either by market sampling or by onboard observers, and processed as described above. For information on recent sampling levels and nations providing samples, see Sec. 2.2. of the most recent HAWG report.
Mean weights-at-age in the stock and proportions mature (maturity ogive) are derived from the June/July international acoustic survey (see next paragraph).

## B.3. Surveys

## B.3.1 Acoustic: ICES Co-ordinated Acoustic Surveys for herring in North Sea, Skagerrak and Kattegat

The ICES Coordinated acoustic surveys started in 1979 around Orkney and Shetland with first major coverage in 1984. An index derived from that survey has been used in assessments since 1994 with the time-series data extending back to 1989. The survey was extended to IIIa to include the overlapping Western Baltic spring spawning stock in 1989, and the index has been used with a number of other tuning indices since 1991. The early survey had occasionally covered VIa (North) during the 1980s and was extended westwards in 1991 to cover the whole of VIa (North). Since 1991, this survey provides the only tuning index for VIa (North) herring and from 2008 for the
whole Malin Shelf, By carrying out the co-ordinated survey at the same time from the Kattegat to Donegall all herring in these areas are covered simultaneously, reducing uncertainly due to area boundaries as well as providing input indices to three distinct stocks. The surveys are co-ordinated under ICES Planning Group for International Pelagic Surveys (PGIPS).

The acoustic recordings are carried out using Simrad EK60 38 kHz sounder echo-integrator with transducers mounted on the hull, drop keel or towed bodies. Prior to 2006, Simrad EK500 and EY500 were also used. Further data analysis is carried out using either BI500, Echoview or Echoann software. The survey track is selected to cover the area giving a basic sampling intensity over the whole area based on the limits of herring densities found in previous years. A transect spacing of 15 nautical miles is used in most parts of the area with the exception of some relatively high density sections, east and west of Shetland, north of Ireland in the Skagerrak where short additional transects were carried out at 7.5 nautical miles spacing, and in the southern area, where a 30 nautical miles transect spacing is used.

The following target strength to fish length relationships have been used to analyse the data:
herring $\quad \mathrm{TS}=20 \log \mathrm{~L}-71.2 \mathrm{~dB}$
sprat

$$
\mathrm{TS}=20 \log \mathrm{~L}-71.2 \mathrm{~dB}
$$

gadoids
$\mathrm{TS}=20 \log \mathrm{~L}-67.5 \mathrm{~dB}$
mackerel $\quad$ TS $=21.7 \log \mathrm{~L}-84.9 \mathrm{~dB}$
Data are reported through standardised data exchange format and uploaded into the FishFrame database, currently held at DTU Aqua, Charlottenlund, Denmark. National estimates are aggregated through Fishframe during PGIPS to calculate global estimates for the North Sea, the Malin Shelf and the western Baltic Sea. The exchange format currently holds information on the ICES statistical rectangle level, with at least one entry for each rectangle covered, but more flexible strata are accommodated by allowing multiple entries for abundance belonging to different strata. Data submitted consists of the ICES rectangle definition, biological stratum, herring abundance by proportion of autumn spawners (North Sea and VIa North) and Spring spawners (Western Baltic, age and maturity, and survey weight (survey track length). Data are presented according to the following age/maturity classes: 1 immature (maturity stage 1 or 2 ), 1 mature (maturity stage $3+$ ), 2 immature, 2 mature, 3 immature, 3 mature, $4,5,6,7,8,9+$. In addition to proportions at age data on mean weights and mean length are reported at age/maturity by biological strata. Data are combined using an effort weighted mean based on survey effort reported as number of nautical miles of cruise track per statistical rectangle. A combined survey report is produced annually. Apart from the Biomass index for 1-9+-ringers, mean weights at age in the catch and proportions mature are derived from the survey to be used in the NSAS assessment.

## B.3.2 International Bottom Trawl Survey:

The International Bottom Trawl Survey (IBTS) started out as a Young Herring Survey (IYHS) in 1966 with the objective of obtaining annual recruitment indices for the combined North Sea herring stocks. It has been carried out every year since, and it was realized that the survey could provide recruitment indices not only for herring, but for roundfish species as well. Examination of the catch data from the 1st quarter IBTS showed that these surveys also gave indications of the abundances of the adult stages of herring, and subsequently the catches have been used for estimating 2-5+
ringer abundances. The surveys are carried out in $1^{\text {st }}$ quarter (February) and in $3^{\text {rd }}$ quarter (August-September) using standardized procedures among all participants. The standard gear is a GOV trawl, and at least two hauls are made in each statistical rectangle. In 2007 the IBTS was extended into English Channel. In addition, historical IBTS indices have been updated from 2004 onwards (in 2007).
In 1977 sampling for late stage herring larvae was introduced at the IBTS $1^{\text {st }}$ quarter, using Isaccs-Kidd Midwater trawls. These catches appeared as a good indicator of herring recruitment, however examination of IKMT performance showed deficiencies in its catchability for herring larvae, and a more applicable gear, a ring net (MIK) was suggested as an alternative gear. Hence, gear type was changed in the mid $90^{\prime}$ ies, and the MIK has been the standard gear of the programme since. This ring net is of 2 meter in diameter, has a long two-legged bridle, and is equipped with a black netting of 1.5 mm mesh size. Two oblique hauls per ICES statistical rectangle are made during night.

Indices of 2-5+ ringer herring abundances in the North Sea (1st quarter). Fishing gear and survey practices were standardised from 1983, and herring abundance estimates of 2-5+ ringers from 1983 onwards has shown the most consistent results in assessments of these age groups. This series is used in North Sea herring assessment. Catches in Division IIIa are not included in this index. These estimates are determined by the standard IBTS methodology developed by the ICES IBTS working group.

Index of 1-ringer recruitment in the North Sea (1 ${ }^{\text {st }}$ quarter). The 1-ringer index of recruitment is based on trawl catches in the entire survey area, hence, all 1-ringer herring caught in Division IIIa is included in this index. Indices are calculated as an area weighted mean over means by ICES statistical rectangle, and are available for year classes 1977 to recent. The Downs herring hatch later than the other autumn spawned herring and generally appears as a smaller sized group during the 1st quarter IBTS. A recruitment index of smaller sized 1-ringers is calculated using the standard procedure, but solely based on abundance estimates of herring <13 cm (ICES CM 2000/ ACFM:10, and ICES CM 2001/ ACFM:12).
MIK index of 0-ringer recruitment in the North Sea (1 ${ }^{\text {st }}$ quarter). The MIK catches of late stage herring larvae are used to calculate an 0-ringer index of autumn spawned herring in the North Sea, this represents recruitment strength (Nash \& Dickey-Collas 2005). A flowmeter at the gear opening is used for estimation of volume filtered by the gear, and using this information together with information on bottom depth, the density of herring larvae per square meter is estimated. The mean herring density in statistical rectangles is raised to mean within subareas, and based on areas of these subareas an index of total abundance is estimated (see also ICES 1996/Asses:10). The series estimates for subareas as well as the total index.

## B.3.3. Larvae:

Surveys of larval herring have a long tradition in the North Sea. Sporadic surveys started around 1880, and available scientific data goes back to the middle of the 20th century. The co-ordination of the International Herring Larvae Surveys in the North Sea and adjacent waters (IHLS) by ICES started in 1967, and from 1972 onwards all relevant data are achieved in a data base (ICES PGIPS). The surveys are carried out annually to map larval distribution and abundance (Schmidt et al., 2009). Larval abundance estimates are of value as relative indicators of the herring spawning biomass in the assessment.

Nearly all countries surrounding the North Sea have participated in the history of the IHLS. Most effort was undertaken by the Netherlands, Germany, Scotland, England, Denmark and Norway. A number of other nations have contributed occasionally. A sharp reduction in ship time and number of participating nations occurred in the end of the 1980s. Since 1994 only the Netherlands and Germany contribute to the larvae surveys, with one exception in 2000 when also Norway participated.

Larvae Abundance Index (LAI): The total area covered by the surveys is divided into 4 sub areas corresponding to the main spawning grounds. These sub areas have to be sampled in different given time intervals. The sampling grid is standardized and stations are approximately 10 nautical miles apart. The standard gear is a GULF III or GULF VII sampler (Nash et al., 1998). Newly hatched larvae less than 10 mm total length ( 11 mm for the Southern North Sea) are used in the index calculation. To estimate larval abundance, the mean number of larvae per square meter obtained from the Ichthyoplankton hauls is raised to rectangles of $30 \times 30$ nautical miles and the corresponding surface area. These values are summed up within the given unit and provide the larval abundance per unit and time interval.

Multiplicative Larval Abundance Index (MLAI): The traditional LAI and LPE (Larval Production Estimates) rely on a complete coverage of the survey area. Due to the substantial decline in ship time and sampling effort since the end of the 80s, these indices could not be calculated in their traditional form since 1994. Instead, a multiplicative model was introduced for calculating a Multiplicative Larvae Abundance Index (MLAI, Patterson \& Beveridge, 1995). In this approach the larvae abundances are calculated for a series of sampling units. The total time series of data are used to estimate the year and sampling unit effects on the abundance values. The unit effects are used to fill unsampled units so that an abundance index can be estimated for each year.

Calculation of the linearised multiplicative model was done using the equation:
$\ln ($ Indexyear,LAI unit $)=$ MLAIyear + MLAILAI unit + uyear, LAI unit
where MLAIyear is the relative spawning stock size in each year, MLAIlai unit are the relative abundances of larvae in each sampling unit and year, LAI unit are the corresponding residuals (Gröger et al., 1999, 2000). The unit effects are converted such that the first sampling unit is used as a reference (Orkney/Shetland 01-15.09.72) and the parameters for the other sampling units are redefined as differences from this reference unit. The model is fitted to abundances of larvae less than 10 mm in length ( 11 mm for SNS). The MLAI is updated annually and represent all larval data since 1972. The time series is used as a biomass index in the herring assessment.

## B.4. Commercial CPUE

Not used for pelagic stocks.

## B.5. Other relevant data

## B.5.1 Separation of North Sea Autumn Spawners and Illa-type Spring Spawners

North Sea Autumn Spawners and IIIa-type Spring Spawners occur in mixtures in fisheries operating in Divisions IIIa and IVaE (ICES, 1991/Assess:15; Clausen et al., 2007): mainly $2+$ ringers of the Western Baltic spring-spawners and 0-2-ringers from the North Sea autumn-spawners, including winter-spawning Downs herring. In addition, several local spawning stocks have been identified with a minor importance for
the herring fisheries (ICES, 2001/ACFM 12).
The method of separating herring in Norwegian samples, using vertebral counts as described in former reports of this Working Group (ICES 1990/ Assess:14) assumes that for autumn spawners, the mean vertebral count is 56.5 and for Spring spawners 55.80. The fractions of spring spawners (fsp) are estimated from the formula (56.50$\mathrm{v}) /(56.5-55.8)$, where v is the mean vertebral count of the (mixed) sample with the restriction that the proportion should be one if $\mathrm{fsp}>=1$ and zero if $\mathrm{fsp}<=0$. The method is quite sensitive to within-stock variation (e.g. between year classes) in mean vertebral counts.

Experience within the Herring Assessment Working Group has shown that separation procedures based on size distributions often will fail. The introduction of otolith microstructure analysis in 1996-97 (Mosegaard \& Popp-Madsen, 1996) enables an accurate and precise split between three groups, autumn, winter and spring-spawners. However, different populations with similar spawning periods are not resolved with the present level of analysis. Different stock components that are not easily distinguished by their otolith microstructure (OM), are considered to have different mean vertebral counts (vs) as, e.g., winter-spawning Downs herring: 56.6 (Hulme, 1995), and the small local stocks, the Skagerrak winter/spring-spawners: 57 (Rosenberg and Palmén, 1982). Further, the estimated stock specific mean vs count varies somewhat among different studies; North Sea: 56.5, Western Baltic Sea: 55.6 (Gröger \& Gröhsler, 2001) and North Sea: 56.5, Western Baltic Sea: 55.8 (ICES 1992/H:5). Comparison between separation methods using frequency distributions of vertebral counts and otolith microstructure showed reasonable correspondence. Using this information the years from 1991 to 1996 was reworked in 2001, applying common splitting keys for all years by using a combination of the vertebral count and otolith microstructure methods (ICES, 2001/ACFM:12). From 2001 and onwards, the otolith-based method only has been used for the Division IIIa.

Different methods of identifying herring stocks in the Division IIIa and Subdivisions 22-24 were evaluated in EU CFP study project (EC study 98/026). The study involved several inter-calibration sessions between microstructure readers in the different laboratories involved with the WBSS herring. After the study was finished a close collaboration concerning reader interpretations has been kept between the Danish and Swedish laboratories. Sub-samples of the 2002 and 2003 Danish, Swedish, and German microstructure analyses were double-checked by the same Danish expert reader for consistency in interpretation. The overall impression is an increasingly good agreement among readers (Clausen et al., 2007).

New molecular genetic approaches for stock separation are being developed within the EU-FP5 project HERGEN (EU project QLRT 200-01370). Sampling of spawning aggregations during spring, autumn and winter has been carried out in 2002 and in 2003 in Division IIIa and in the Western Baltic at more than 10 different locations. Preliminary results point at a substantial genetic variation between North Sea and Western Baltic herring (Bekkevold et al., 2005; 2007; Ruzzante et al., 2006) .

After the introduction of otolith microstructure analysis in 1996 it was discovered that in the western Baltic a small percentage of the herring landings might consist of au-tumn-spawners individuals. Before molecular genetic methods became available for Atlantic herring the existence of varying proportions of autumn spawners in Subdivisions 22-24 in different years was considered a potential problem for the assessment.

## C. Historical Stock Development

## C. 1 Model used:

A benchmark assessment for North Sea herring was carried out in 2006. Following the benchmark investigation in 2006, the tool for the assessment of North Sea herring is ICA. However, the environment to execute the ICA has changed from the original ICA software into FLR (now called FLICA). Justification of the choice of assessment model, catch and survey weightings and the length of separable period are found in HAWG 2006 and Simmonds (2003; 2009). After extensive testing HAWG assumes there are no differences between the old ICA and FLICA. Thus FLICA was used to carry out the assessments after 2008.

The assessment has the same set-up and basic assumption as the assessment that was carried out last year. Input data are given in Tables 2.6.2.2. The ICA programme operates by minimising the following general objective function:
$\sum \lambda_{c}(C-\hat{C})^{2}+\sum \lambda_{i}(I-\hat{I})^{2}+\sum \lambda_{r}(R-\hat{R})^{2}$
which is the sum of the squared differences for the catches (separable model), the indices (catchability model) and the stock-recruitment model.

The final objective function chosen for the stock assessment model was:

$$
\begin{aligned}
& \sum_{a=0, y=1997}^{a=8, y=2022} \lambda_{a}\left(\ln \left(\hat{C}_{a, y}\right)-\ln \left(C_{a, y}\right)\right)^{2}+ \\
& \sum_{y=1979}^{y=2002} \lambda_{\text {mlai }} \cdot\left(\ln \left(q_{\text {mlai }} \cdot S \hat{S} B_{y}^{K}\right)-\ln \left(M L A I_{y}\right)\right)^{2}+ \\
& \sum_{a=1, y=1983 * *}^{a=5+y=2003} \lambda_{a, b \text { bbsa }}\left(\ln \left(q_{a, \text { ibsta }} \cdot \hat{N}_{a, y}\right)-\ln \left(\text { IBTS }_{a, y}\right)\right)^{2}+ \\
& \sum_{a=1, y=1989}^{a=9+y=2002} \lambda_{a, a c o u s t}\left(\ln \left(q_{a, a c o u s t} \cdot \hat{N}_{a, y}\right)-\ln \left(\text { ACOUST }_{a, y}\right)\right)^{2}+ \\
& \sum_{y=1977}^{y=2003} \lambda_{\text {mik }}\left(\ln \left(q_{\text {mik }} \cdot \hat{N}_{o, y}\right)-\ln \left(M I K_{y}\right)\right)^{2}+ \\
& \sum_{y=1960}^{y=2002} \lambda_{\text {ssr }}\left(\ln \left(\hat{N}_{o, y+1}\right)-\ln \left(\frac{\alpha S \hat{S} B_{y}}{\beta+S \hat{S} B_{y}}\right)\right)^{2}
\end{aligned}
$$

** except for 1 ring IBTS which runs from 1979 to 2002
with the following variables:

| a,y | age (rings) and year <br> C |
| :--- | :--- |
| $\hat{C}$ | Estimated at age (rings) |
| $\hat{N}$ | Estimated population numbers |
| $\hat{S S B}$ | Estimated spawning stock size |
| MLAI | MLAI index (biomass index) |
| ACOUST | Acoustic index (age disaggregated) <br> IBTS |
| IBTS index (1-5+ ringers) |  |
| q | MIK index (0-ringers) |
| k | Catchability |
| $\alpha, \beta$ | power of catchability model |
| parameters to the Beverton stock-recruit model |  |

$\lambda \quad$ Weighting factor
Software used: FLICA, based on ICA (Patterson, 1998; Needle, 2000; Kell et al., 2007)
Model Options chosen:
The model settings should be as follows (as determined by the last benchmark, HAWG 2006)

| FLICA control settings | Settings | Description |
| :--- | :--- | :--- |
| sr | TRUE | Stock and recruitment <br> relationship |
| sr.age | 1 | age at recruitment |
| lambda.age | 0.10 .13 .672 .872 .231 .741 .37 | Weighting matrices for catch- <br> at-age; for aged surveys; for <br> SSB surveys |
| lambda.yr | 1.040 .940 | Relative weights by year |
| lambda.sr | 0.1 | weight for the SRR term in <br> the objective function |
| index.model | pawer linear linear linear | Catchability model for each <br> survey |
| index.cor | 5 | Are the age-structured <br> indices are correlated across <br> ages |
| sep.nyr | 4 | Number of years for <br> separable model |
| sep.age | 1 | Reference age for fitting the <br> separable model |
| sep.sel | Selection on last true <br> reference age |  |

Input data types and characteristics:

| TYPE | NAME | YEAR <br> RANGE | AGE <br> RANGE | VARIABLE FROM <br> YEAR TO YEAR <br> YES/NO |
| :--- | :--- | :--- | :--- | :--- |
| Caton | Catch in tonnes |  |  |  |
| Canum | Catch at age in numbers | $1960-2008$ | $1-9+$ | Yes |
| Weca | Weight at age in the commercial catch | $1960-2008$ | $1-9+$ | Yes |
| West | Weight at age of the spawning stock at <br> spawning time. | $1960-2008$ | $1-9+$ | Yes (3 year running <br> mean) |
| Mprop | Proportion of natural mortality before <br> spawning | $1960-2008$ | $1-9+$ | No |
| Fprop | Proportion of fishing mortality before <br> spawning | $1960-2008$ | $1-9+$ | No |
| Matprop | Proportion mature at age | $1960-2008$ | $1-9+$ | Yes |
| Natmor | Natural mortality | $1960-2008$ | $1-9+$ | No |

Tuning data:

| TYPE | NAME | YEAR RANGE | AGE RANGE (WR) |
| :--- | :--- | :--- | :--- |
| Tuning fleet $\mathbf{1}$ | IBTS Q1 | $1984-2009$ | $1-5$ |
| Tuning fleet $\mathbf{2}$ | MIK | $1992-2009$ | 0 |
| Tuning fleet 3 | Acoustic | $1989-2008$ | $1-9+$ |
| Tuning fleet 4 | MLAI | $1973-2008$ | SSB |

## C. 2 Variance and weighting factors for ICA

In the ICA model a fixed set of inverse variance weights for surveys and catch at age have been used. In the benchmark assessment in 2006 (ICES 2006/ACFM:20) the weighting factors of the indices used in ICA were fixed and have been used with the same values since. This reflects a slight change from a major investigation in 2001 carried out by the Study Group on Evaluation of Current Assessment Procedures for North Sea herring (SGEHAP, ICES 2001/ACFM:22). The original weighting factors were derived from the survey and catch data by methods given in ICES 2001/ACFM:22 and Simmonds (2003). The variance used is the variance of the natural logarithm of the estimates of the index based on a 2 stage bootstrap procedure. The choice matches the use of a maximum log likelihood method with a lognormal error distribution used within the ICA model. All indices are treated in the same manner. The individual station estimates at all ages are bootstrapped using a simple resampling with replacement procedure. This provides a variance covariance estimate of estimates of indices at age for each index assuming identically independently distributed samples. (iid)

As the spatial distributions are correlated and the sampling on the surveys are nonrandom in space, the spatial autocorrelation was taken into account using geostatistics. The methodology is described in Rivoirard et al. (2000), who provide the formulae and methods required to estimate variograms and calculate the estimation variance. Petitgas and Lafont (1997) provide the free software (EVA2) that has been used here for calculating the estimation variance for all the surveys. The iid estimates are corrected to provide overall estimates of variance covariance estimates across ages for each survey. The mean variance covariance estimate for the survey timeseries was calculated to provide one average variance/covariance matrix per survey.

ICA does not explicitly deal with covariance (in common with many assessment models) but it does allow modification of weights at age to account for this in a general way. The concept is to reduce the inverse variance factor by an amount that accommodates the covariance. The limits are: for zero correlation a factor of unity; for $100 \%$ covariance over n ages weights of $1 / \mathrm{n}$. In both surveys the 1 to 2 group estimates are effectively independent and can be given weighting due to the full inverse variance weight, for subsequent ages the weighting has been implemented here for intermediate values of covariance to give the Wage weighting factors at age:

$$
W_{\text {age }}=\frac{1}{\operatorname{var}_{\text {age }}}\left\{n-\sum \operatorname{cov}_{\text {age,age-1 }}\right\} /\left\{\operatorname{cov}_{\text {age,age-1 }} / \sum 1 / \operatorname{cov}_{\text {age,age-1 }}\right\}
$$

Where varage is the variance of $\ln$ (estimate at age)
cov is covariance (age, age- 1 )
$n$ is the number of ages in the correlated sequence

The resulting correlation correction factors are given in Table 2.6.7.3 in HAWG Report 2008.

The weighting factors used since 2006 (ICES 2006/ACFM:20) are given in Table 1 and can be compared with the old weighting factors derived under SGEHAP (ICES 2001/ACFM:22). The major difference is a slight general reduction in survey weights relative to the catch. Among the surveys the resulting spread of weights is generally similar to the earlier values, reducing with age, more steeply with the IBTS than the acoustic. The major difference is the MIK weighting which is reduced to about $1 / 3$ of the previous value. The change is caused by the recent extended analysis. The difference between the previous analysis and this one was that in the earlier work the geostatistical analysis of spatial variance was limited to only a few recent years in each series. This resulted quite accidentally and unknowingly in selecting years from the MIK index that were very precise.

Table 1: North Sea herring. New weighting factors (ICES 2006 /ACFM:20) based on bootstrap of survey data. Old weights are included for comparison

|  | Catch | Acoustic |  |  | IBTS |  | MIK |  | MLAI |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | Old | New | Old | New | Old | New | Old | New | Old | New |
| 0 | 0.10 | 0.10 |  |  |  |  | 2.05 | 0.63 |  |  |
| 1 | 0.10 | 0.10 | 0.74 | 0.63 | 0.67 | 0.47 |  |  |  |  |
| 2 | 3.17 | 3.67 | 0.75 | 0.62 | 0.24 | 0.28 |  |  |  |  |
| 3 | 2.65 | 2.87 | 0.64 | 0.17 | 0.06 | 0.01 |  |  |  |  |
| 4 | 1.94 | 2.23 | 0.27 | 0.10 | 0.03 | 0.01 |  |  |  |  |
| 5 | 1.31 | 1.74 | 0.14 | 0.09 | 0.03 | 0.01 |  |  |  |  |
| 6 | 0.97 | 1.37 | 0.13 | 0.08 |  |  |  |  |  |  |
| 7 | 0.75 | 1.04 | 0.12 | 0.07 |  |  |  |  |  |  |
| 8 | 0.55 | 0.94 | 0.07 | 0.07 |  |  |  |  |  |  |
| 9 | 0.54 | 0.91 | 0.07 | 0.05 |  |  |  |  |  |  |
| $S S B$ |  |  |  |  |  |  |  |  |  |  |

## D. Short-Term Projection

The short-term prediction method was substantially modified in 2002. Following the review by SGEHAP (ICES 2001/ACFM:22), which recommended that a simple multifleet method would be preferable, the complex split-factor method used for a number of years prior to 2002 has not been used since. The multi-fleet, multi-option, deterministic short-term prediction programme (MFSP) was accepted by ACFM in 2002 and further refined in 2003. It has been used routinely to perform short term predictions for this stock since then.. The good agreement between predicted biomass for the acintermediate year and SSB taken from the assessment one year after demonstrates that the current prediction procedure for stock numbers is working well.

## Method

The procedure and programme used (MFSP Skagen; WD to HAWG 2003) was the same as has been used since 2003. For the North Sea herring, managers have agreed to constrain the total outtake at levels of fishing mortalities for ages $0-1$ and 2-6, and need options to show the trade-off between fleets within those limits. The MFSP program was developed to cover these needs.

## Input data

## Fleet Definitions

The current fleet definitions are:

## North Sea

Fleet A: Directed herring fisheries with purse seiners and trawlers. By-catches in industrial fisheries by Norway are included.

Fleet B: Herring taken as by-catch under EU regulations.

## Division IIIa

Fleet C: Directed herring fisheries with purse seiners and trawlers
Fleet D: By-catches of herring caught in the small-mesh fisheries
The fleet definitions are the same as last year.
In some years, it has been agreed that Norway can take parts of its IIIa quota in the North Sea. When estimating the expected catch in the intermediate year, it is assumed that this transfer takes place, hence the assumed catch by the C-fleet of both stocks combined is reduced and the catch by the A-fleet increased with the agreed amount.

Input Data for Short Term Projections: All the input data for the short term projections are shown in Table 2.7.1, which is the input file for the predictions.

Stock Numbers: For the start of the intermediate year the stock numbers at age by 1. Jan that year are taken from the prediction made by ICA.

Recruitment: For the prediction years, the recruitment has in recent years been set to the geometric mean of the recruitments of the year classes from 2001 onwards, as estimated in this year's assessment. The low recruitment was assumed because all the year classes from 2001 onwards have been poor except for 2008 year class. Analysis of the time series of SSB and recruitment data by the SGRECVAP (ICES CM 2006/LRC:03) clearly indicates a shift in the recruitment success in 2001. The underlying cause for the change in 2001 is not clear, but there is no evidence to justify an assumption of long term average recruitment in the near future. Consequently, the advice is adapted to the current low recruitment regime.

Fishing Mortalities: Selection by fleet at age is calculated by splitting the total fishing mortality in the last assessment year at each age (from the assessment output) proportional to the catches by fleets at that age. These selections at age were used for all years in the prediction.

Mean weights in the catch by fleet: The 3 year average mean weights at age for each fleet are used for all prediction years, unless there are indications that some year class has abnormal growth.

Mean Weights at age in the stock: The weights at age applied in the last assessment year were used for all predictions years. These are running averages of the raw data. In previous years, the procedure was different, to account for the special growth of the 2000 year class.

Maturity at age: The 3 years average maturity was used.
Natural Mortality: Equal to those assumed in the assessment.
Proportion of M and F before spawning: Standard values of 0.67 for both

## Prediction

## Assumptions for the intermediate year.

A-fleet: The TAC for the A fleet has been over-fished every year since 2003, and it is assumed that this will be the case in the intermediate year as well. Unless there are strong indications of a change in practise, the percentage assumed is the average over the last 3 years.

The catches by the B-fleet have been well below the by-catch quota for the B-fleet. The quota has been reduced recently, and the fraction used has increased. Therefore, the same fraction as last year is assumed. Also the C and D fleets have catches well below the quota, partly because the quota also includes WBSS herring. For 2009, the same fraction as in 2008 was assumed; previously a 3 year average has been used in some cases.

Points of interpretation:

- In years when Norway is allowed to transfer some of its quota in IIIa to IV, this transfer is assumed in the predictions


## Management Option Tables for the TAC year

The EU-Norway agreement on management of North Sea herring was updated in 2008, to adapt to the present reduced recruitment, accounting for the results by WKHMP. The revised rule specifies fishing mortalities for juveniles ( $\mathrm{F}_{0-1}$ ) and for adults ( $\mathrm{F}_{2-6}$ ) not to be exceeded, at 0.05 and 0.25 respectively, for the situation where the SSB is above 1.5 million tonnes. When the SSB is below 1.5 million tonnes. Moreover, the current agreement has a constraint on year-to-year change of $15 \%$ in TAC, F is reduced to give
$\mathrm{F}_{2-6}=0.25-\left(0.15^{*}(1500-S S B) / 700\right)$,
with allowance for a stronger reduction in TAC if necessary.
Furthermore, there is a constraint at $15 \%$ change in the TAC from one year to the next.

- The $\mathrm{F}_{0-1}$ and $\mathrm{F}_{2-6}$ stated in the rule are assumed to apply to the total F summed over all fleets.
- The SSB referred to is taken to be the SSB in the prediction year, i.e. the fishing mortalities for 2010 should reflect its consequence for SSB in 2010.

Catches by the C and D fleet influence the fishing opportunities for the B-fleet in particular, since the NSAS herring caught by these fleets mostly are at age 0-2. The assumed catch of NSAS herring by the C and D fleets is derived according to a likely TAC for WBSS herring in a three step procedure:

1. The fraction of the total TAC for WBSS that is taken in Division IIIa is assumed to be the same as last year, giving an expected catch of WBSS in Division IIIa.
2. The WBSS caught in Division IIIa is allocated to the $C$ and D fleets assuming the same share as last year. The total expected catch of WBSS in IIIa is split accordingly, which gives expected catch of WBSS by fleet.
3. Using the ratio between NSAS and WBSS in the catches by each fleet, the total catch by fleet and the catch of NSAS by fleet are derived from the catch of WBSS by fleet.

These expected catches of NSAS by the C and D fleets are used as catch constraints in
the prediction.
The basis for deriving these catches is weak. The main purpose is to provide realistic assumptions on the impact of these fleets when predicting the catches for the North Sea fleets. The effect of other assumptions for the $C$ and $D$ fleet should be calculated if needed, but are not presented in the advice.

The catches for the A and B fleets are derived according to the harvest rule.
When the harvest rule leads to SSB below the trigger biomass ( 1.5 million tonnes), an iterative procedure is needed to find a fishing mortality and a corresponding SSB in accordance with the rule. At present, this is done manually by scanning over ranges of F for the A and B fleet.

## E. Medium-Term Projections - -are made as needed.

Model used: 10 year stochastic prediction Software used: STPR3 has been used as a standard in the past, as it allows for independent regulations of two 'flles' (fisheries)

Initial stock size: As for the short term prediction, but with random variation according the variance-covariance matrix taken from the ICA assessment

Natural mortality: Constant as in the assessment
Maturity: As in the short term prediction
F and M before spawning: Constant values : 0.67 for both.
Weight at age in the stock: Obtained each projection year by drawing a historical year randomly and using the weights from that year.

Weight at age in the catch: As weight at age in the stock.
Exploitation pattern: As for short term forecast. Fleet A separately, fleets B-C-D merged.

Intermediate year assumptions: As for short term prediction
Stock recruitment model used: Beverton Holt or Hockey stick
Uncertainty models used:
Initial stock size: See above
Natural mortality: Constant
Maturity: Constant
F and M before spawning: Constant
Weight at age in the stock: See above
Weight at age in the catch: See above
Exploitation pattern: Constant
Intermediate year assumptions: Constant
Stock recruitment model used: Log-normal variation around a stock-recruit function with fixed parameters. Opportunity to truncate the distribution.
F. Long-Term Projections - -not done since 1996(?)

## G. Biological Reference Points

The precautionary reference points for this stock were adopted in 1998. The situation
has now arisen that North Sea herring is nominally being managed by a precautionary management plan, although the SSB is now below the precautionary biomass reference point. We consider that the critical issue is identifying the risk of SSB falling below Blim. The following section is adapted from ICES WKHMP (ICES CM 2008 (ACOM:27)) and explores and discusses the issues about precautionary status of the management of North Sea herring.

## The Blim

The 1998 Study Group on Precautionary Approach to Fisheries Management (ICES CM 1998/ACFM:10.) determined reference points for North Sea herring that were adopted by ACFM (ICES CM 1998/ACFM:10.). The Blim (800 000 tonnes) was set at a level below which the recruitment may become impaired and was also the formally used MBAL. In 2007, WKREF (ICES CM 2007/ACFM:05) explored limit reference points for North Sea herring and concluded that there is no basis for changing Blim. A low risk of SSB falling below Blim is therefore the basis of ICES precautionary advice.

## Fpa and Bpa

The target and trigger points used in the management plan (which began in 1997) were recommended by the Study Group on Precautionary Approach to Fisheries Management and adopted by ACFM as the precautionary reference points. This means that the precautionary reference points were taken from the already existing management plan. In the management plan, the target fishing mortalities were intended as targets and not as bounds. The higher inflection point (B trigger) in the earlier rule ( 1.3 million tonnes) was derived largely as a compromise, allowing higher exploitation at higher biomass but reflecting an ambition to maintain the stock at a high level, by reducing the fishing mortality at an early stage of decline. This trigger was changed in November 2008 to 1.5 million tonnes after WKHMP and consultation with the stakeholders. Thus currently the trigger and Bpa are different at 1.5 million tonnes and 1.3 million tonnes respectively.

## Concept of a management plan (harvest control rule)

In a harvest control rule, parameters (trigger and targets) serve as guidance to actions according to the state of the stock (ICES Study Group on the Precautionary Approach, ICES CM 2002/ACFM:10). These should be chosen according to management objectives, one of which should be to have a low risk of bringing the SSB to unacceptably low levels. In the evaluation of a harvest rule, one will use simulations with a 'virtual stock' which as far as possible resembles the stock in question, and the risk is evaluated as the probability of the virtual SSB being below the Blim value. Within the constraints needed to keep the risk to Blim low, parameters of the rule will be chosen to serve other management objectives, e.g. to ensure a high long term yield and stable catches over time. Such a management plan would be classed by ICES as precautionary provided the risk of SSB being below Blim is sufficiently low.

## Concept of precautionary reference points

Conceptually, precautionary reference points (Bpa) are different from parameters in a harvest control rule. In the precautionary approach, as interpreted by ICES, the function of the reference points is to ensure that the SSB is above the range where recruitment may be impaired or the stock dynamics is unknown. The real limit is represented by Blim, while the Bpa takes assessment uncertainty into account, so that
if SSB is estimated at Bpa, the probability that it is below Blim shall be small. The Flim is the fishing mortality that corresponds to Blim in a deterministic equilibrium. The Fpa is related to Flim the same way as Bpa is related to Blim (ICES Study Group on the Precautionary Approach 2002b). In the advisory practice, Fpa has been the basis for the advice unless the SSB has been below Bpa, where a reduction in F has been advised. Furthermore, Fpa and Bpa are currently used to classify the state of stock and rate of exploitation relative to precautionary limits. Precautionary reference points are used by ICES to provide advice and classify the state of the stock in the absence of other information, such as extensive evaluations of management plans.

ICES will accept that a harvest control rule is in accordance with the precautionary approach as long as it implies a low risk to being below Blim, even if other reference points may be exceeded occasionally. When a rule is regarded as precautionary, ICES gives its advice according to the rule. If the rule is followed, then ICES classifies exploitation as precautionary. Within this framework, other precautionary reference points generally will be redundant. However, the precautionary reference points may also be used to classify the stock with respect to precautionary limits, which may lead to a conflicting classification. This discrepancy is still unresolved. For North Sea herring in the present situation, with a reduced recruitment, the SSB may be expected to be below 1.3 million tonnes most of the time. The management plan will reduce fishing mortality accordingly. Following the acceptance by ACFM that the management plan is precautionary (and the findings of WKHMP), HAWG considers that the parameters of the management plan should take primacy over the management against precautionary reference points Fpa or Bpa.

## H. Other Issues

## H. 1 Biology of the species in the distribution area

The herring (Clupea harengus) is a pelagic species which is widespread in its distribution throughout the North Sea. The herring's unique habit is that it produces benthic eggs which are attached to a gravely substrate on the seabed (Geffen 2009). This points strongly to an evolutionary history in which herring spawned in rivers and at some later date re-adapted to the marine environment(Geffen 2009). The spawning grounds in the southern North Sea are in fact located in the beds of rivers which existed in geological times and some groups of spring spawning herring still spawn in very shallow inshore waters and estuaries. Spawning typically occurs on coarse gravel $(0.5-5 \mathrm{~cm})$ to stone $(8-15 \mathrm{~cm})$ substrates and often on the crest of a ridge rather than hollows. For example, in a spawning area in the English Channel, eggs were found attached to flints $2.5-25 \mathrm{~cm}$ in length, where these occurred in gravel, over a 3.5 km by 400 m wide strip.

As a consequence of the requirement for a very specific substrate, spawning occurs in small discrete areas in the near coastal waters of the western North Sea (Schmidt et al., 2009). They extend from the Shetland Isles in the north through into the English Channel in the south. Within these specific areas actual patches of spawn can be extremely difficult to find.

The fecundity of herring is length related and varies between approximately 10000 and 60000 eggs per female (Damme et al in press). This is a relatively low fecundity for teleosts, probably because. The age of first maturity is 3 years old (2 ringers) but the proportion mature at age may vary from year to year dependent on feeding conditions. Over the past 15 years the proportion mature at age 3 years (2 ringers) has
ranged from $47 \%$ to $86 \%$ and for 4 year old fish ( 3 winter ringers) from $63 \%$ to $100 \%$. Above that age, all are considered to be mature.

The benthic eggs take about three weeks to hatch dependant on the temperature. The larvae on hatching are 6 mm to 9 mm long and rise due to buoyancy changes to become planktonic (Dickey-Collas et al, 2009). Their yolk sac lasts for a few days during which time they will begin to feed on phytoplankton and small zooplankton. Their planktonic development lasts around three to four months during which time they are passively subjected to the residual drift which takes them to various coastal nursery areas on both sides of the North Sea and into the Skagerrak and Kattegat (Heath et al., 1997).

Herring continue to be mainly planktonic feeders throughout their life history although there are numerous records of them taking small fish, such as sprat and sandeels, on an opportunistic basis. Calanoid copepods, such as Calanus, Pseudocalanus and Temora and the Euphausids, Meganyctiphanes and Thysanoessa still form the major part of their diet during the spring and summer (Hardy, 1924; Savage, 1937; Bainbridge and Forsyth, 1972; Last, 1989) and are responsible for the very high fat content of the fish at this time. They also consume fish eggs (Segers et al., 2007).

In the past, herring age has been determined by using the annual rings on the scales. In more recent years the growth rings on the otolith have proved more reliable for age determination. Herring age is expressed as number of winter rings on the otolith rather than age in years as for most other teleost species where a nominal 1 January birthdate is applied. Autumn spawning herring do not lay down a winter ring during their first winter and therefore remain as ' 0 ' winter ringers until the following winter. When looking at year classes, or year of hatching, it must be remembered that they were spawned in the year prior to their classification as ' 0 ' winter ringers.

North Sea herring comprise both spring and autumn spawning groups, but the major fisheries are carried out on the offshore autumn spawning fish. The spring spawners are found mainly as small discrete coastal groups in areas such as The Wash, the Thames estuary, Danish Fjords and the now extinct Zuiderzee herring. Juveniles of the spring spawning stocks are found in the Baltic, Skagerrak and Kattegat, and may also be found in the North Sea as well as Norwegian coastal spring spawners.

The main autumn spawning begins in the northern North Sea in August and progresses steadily southwards through September and October in the central North Sea to November and as late as January in the southern North Sea and eastern English Channel. The widespread but discrete location of the herring spawning grounds throughout the western North Sea has been well known and described since the $19^{\text {th }}$ century (Heincke, 1898; Bjerkan, 1917). This led to considerable scientific debate and eventually to investigation and research on stock identity. The controversy centred on whether or not the separate spawning grounds represented discrete stocks or 'races' within the North Sea autumn spawning herring complex (McQuinn, 1997). Resolution of this issue became more urgent as the need for the introduction of management measures increased during the 1950's. The International Council for the Exploration of the Sea (ICES) encouraged tagging and other racial studies and a review of all the historic evidence to resolve this problem and innovative approaches to assessing mixed and connective stocks (Secor et al., 2009; Kell et al., 2009). The conclusions were the basis for establishing the working hypothesis that the North Sea autumn spawning herring comprise a complex of at least four spawning components each with separate spawning grounds, migration routes and nursery areas. There is mixing between these components during the summer

The main four spawning components are:

- The Orkney/Shetland component which spawn from July to early September in the Orkney Shetland area. Nursery areas for fish up to two years old are found along the east coast of Scotland and also across the North Sea and into the Skagerrak and Kattegat.
- The Buchan component which spawn from August to early September off the Scottish east coast. Nursery areas for fish up to two years old are found along the east coast of Scotland and also across the North Sea and into the Skagerrak and Kattegat.
- The Banks or central North Sea component, which derive their name from their former spawning grounds around the western edge of the Dogger Bank. These spawning grounds have now all but disappeared and spawning is confined to small areas along the English east coast, from the Farne Islands to the Dowsing area, from August to October. The juveniles are found along the east coast of England, down to the Wash, and also off the west coast of Denmark.
- The Downs component which spawns in very late Autumn through to February in the southern Bight of the North Sea and in the eastern English Channel. The drift of their larvae takes them north-eastwards to nursery areas along the Dutch coast and into the German Bight (Burd 1985).

At certain times of the year, individuals from the three stock units may mix and are caught together as juveniles and adults but they cannot be readily separated in the commercial catches other than using otolith methods (Clausen et al, 2007). However North Sea autumn spawning herring are managed as a single unit with the understanding that they comprise of many spawning components.

A further complication is that juveniles of the North Sea stocks are found, outside the North Sea, in the Skagerrak and Kattegat areas and are caught in various fisheries there. The proportions of juveniles of North Sea origin, found in these areas varies with the strength of the year class, with higher proportions in the Skagerrak and Kattegat when the year class is good.

Recruitment strength is determined during the larval phase (Nash \& Dickey-Collas 2005) and this is likely to occur prior to the larvae being 20mm in length (Oeberst et al., 2009).

## H. 2 Historic stock development and history of the fishery

Over many centuries the North Sea herring fishery has been a cause of international conflict sometimes resulting in war, but in more recent times in bitter political argument. There have also been fundamental changes in the nature of the fisheries (Poulsen, 2006). These have been driven both by changes in catching power and in response to changes in market requirements, particularly the demand for fish meal and oil. Most of these changes have resulted in greater exploitation pressures that increasingly led to the urgent need to ensure a more sustainable exploitation of North Sea herring. Such pressures really began to exert themselves for the first time during the 1950's when the spawning stock biomass of North Sea autumn spawning herring fell from 5 million tonnes in 1947 to 1.4 million tonnes by 1957 (Simmonds 2007, 2009). That period also witnessed the decline and eventual disappearance of a traditional autumn drift net fishery in the southern North Sea (Burd, 1978).

The annual landings from 1947 through to the early 1960's were high, but stable, av-
eraging around 650000 t (Cushing and Bridger, 1966). Over the period 1952-62, the high fishing mortality (F 0.4 ages 2-6) resulted in a rapid decline in the spawning stock biomass from around 5 million tonnes to 1.5 million tonnes. Recruitment over this period was reasonable, but there were fewer and fewer year classes present in the adult stock, a clear indication that the stocks were being over-fished and that they were also being impacted by the developing industrial fishery in the eastern North Sea.

This period witnessed the complete collapse of the historic East Anglian autumn drift net fishery, which was based entirely on the Downs stock moving south to the Southern Bight and eastern English Channel to spawn. The reasons for that failure have been attributed both to high mortality of the juveniles in the North Sea industrial fisheries, and to heavy fishing by bottom trawlers on the spawning concentrations, in the English Channel, during the 1950's. Such intensive trawling, on vulnerable spawning fish, not only generated a high mortality but also disturbed spawning aggregations, destroyed the spawn and damaged the substrate on which successful spawning depends.

Fishing mortality on the herring in the central and northern North Sea began to increase rapidly in the late 1960's and had increased to F1.3 ages $2-6$, or over $70 \%$ per year of those age classes, by 1968. Landings peaked at over 1 million tonnes in 1965, around $80 \%$ of which were juvenile fish. This was followed by a very rapid decline in the SSB and the total landings. By 1975 the SSB had fallen to 83500 t , although the total landings were still over $300000 t$ (Simmonds 2007). At the same time, spawning in the central North Sea had contracted to the grounds off the east coast of England whilst spawning grounds around the edge of the Dogger Bank were no longer used. This heralded the serious decline and near collapse of the North Sea autumn spawning herring stock which led to the moratorium on directed herring fishing in the North Sea from 1977 to 1981 (Cushing, 1992).
International larvae surveys and acoustic surveys were used to monitor the state of the stocks during the moratorium. By 1980 these surveys were indicating a modest recovery in the SSB from its 1977 low point of 52000 t . By 1981 the SSB had increased to over 200000 t . Prior to the moratorium there had been no control, other than market forces, on catches in the North Sea directed herring fishery. Once the fishery reopened in 1981 the North Sea autumn spawning herring stock was managed by a Total Allowable Catch (TAC) constraint. It should be noted that the TAC was only applied to the directed herring fishery in the North Sea which exploited mainly adult fish for human consumption. Targeted fishing for herring for industrial purposes was banned in the North Sea in 1976 but there was a $10 \%$ by-catch allowance in the fisheries for other species, including the small meshed fisheries for industrial purposes, mainly for sprat. Following the re-opening of the now controlled fishery the SSB steadily increased, peaking at 1.3 million tonnes in 1989. Annual recruitment, measured as ' 0 'group fish, was well above the long-term average over this period. The 1985 year class was the biggest recorded since 1960 and the third highest in the records dating back to 1946. Landings also steadily increased over this period reaching a peak of 876000 tonnes in 1988. This resulted from a steady increase in fishing mortality to $\mathrm{F}_{\text {ages } 2-6}=0.6$ (ca. $45 \%$ ) in 1985 and a high by-catch of juveniles in the industrial fisheries for sprat. Following a period of four years of below average recruitment (year classes 1987-91), SSB fell rapidly to below 500000 tonnes in 1993. Fishing mortality increased rapidly averaging $\mathrm{F}_{\text {ages }} 2-6=0.75$ (ca. $52 \%$ ) over the period 1992-95 and recorded landings regularly exceeded the TAC. The North Sea industrial fishery for sprat developed rapidly over this period with the annual catch increasing from 33000
tonnes in 1987 to 357000 tonnes by 1995. With the $10 \%$ by-catch limit as the only control on the catch of immature herring, there was a consequent high mortality on juvenile herring which averaged $76 \%$ of the total catch in numbers of North Sea autumn spawners over this period.

During the summer of 1991 the presence of the parasitic fungus Ichthyophonus spp was noted in the North Sea herring stock. All the evidence suggested that the parasite was lethal to herring and that its occurrence could have a significant effect on natural mortality in the stock and ultimately on spawning stock biomass. High levels of infection were recorded in the northern North Sea north of latitude $60^{\circ} \mathrm{N}$ whilst infection rates in the southern North Sea and English Channel were very low. Efforts were made to estimate the prevalence of the disease in the stock through a programmeme of research vessel and commercial catch sampling. This led to estimates of annual mortality up to $16 \%$ (Anon., 1993) which was of the same order as the estimate of fishing mortality at the time. It was recognised that the behavioural changes and catchability of infected fish affected the reliability of the estimate of prevalence of the disease in the population. The uncertainty about the effect on stock size varied between estimates of $5 \%$ to $10 \%$ and $20 \%$. Continued monitoring of the progress of the disease showed that by 1994 the prevalence in the northern North Sea had fallen from $5 \%$ in 1992 to below $1 \%$ and confirmed that the infection did not appear to be spreading to younger fish. Ultimately it was concluded that the disease had caused high mortality in the northern North Sea during 1991 and subsequently declined to the point where by 1995 the disease induced increase in natural mortality was insignificant.

The increased fishing pressure during the first half of the 1990's and the disease induced increase in natural mortality led to serious concerns about the possibilities of a stock collapse similar to that in the late 1970's. Reported landings continued at around 650000 tonnes per year whilst the spawning stock began to decline again from over 1 million tonnes in 1990. The assessments at that time were providing an over optimistic perception of the size of the spawning stock and, for example, it was not until 1995 that it was realised that the SSB in 1993 had already fallen below 500 000 tonnes. This was well below the minimum biologically accepted level of 800000 tonnes (MBAL) which had been set for this stock at that time.

## H. 3 Management and ICES advice

In 1996, the total allowable catches (TACs) for Herring caught in the North Sea (ICES areas IV and Division VIId) were changed mid-year with the intention of reducing the fishing mortality by $50 \%$ for the adult part of the stock and by $75 \%$ for the juveniles. For 1997, the regulations were altered again to reduce the fishing mortality on the adult stock to 0.25 and for juveniles to less than 0.1 with the aim of rebuilding the SSB up to 1.1 million t in 1998 (Simmonds 2007).

According to the EU and Norway agreement adopted in December 1997, efforts should be made to maintain the SSB above the MBAL (Minimum Biologically Acceptable Level) of 800000 tonnes. An SSB reference point of 1.3 million has been set above which the TACs will be based on an $\mathrm{F}=0.25$ for adult herring and $\mathrm{F}=0.12$ for juveniles. If the SSB falls below 1.3 million tonnes, other measures will be agreed and implemented taking account of scientific advice. The management agreement was revised in 2004 and now reads:

The stock is managed according to the EU-Norway Management agreement which was updated inNovember 2008, the relevant parts of the text are included here for reference:

1. Every effort shall be made to maintain a minimum level of Spawning Stock Biomass (SSB) greater than 800,000 tonnes (Blim).
2. Where the SSB is estimated to be above 1.5 million tonnes the Parties agree to set quotas for the directed fishery and for bycatches in other fisheries, reflecting a fishing mortality rate of no more than 0.25 for 2 ringers and older and no more than 0.05 for $0-1$ ringers.
3. Where the SSB is estimated to be below 1.5 million tonnes but above 800,000 tonnes, the Parties agree to set quotas for the direct fishery and for bycatches in other fisheries, reflecting a fishing mortality rate on 2 ringers and older equal to:
$0.25-\left(0.15^{*}(1,500,000-S S B) / 700,000\right)$ for 2 ringers and older, and no more than 0.05 for $0-1$ ringers
4. Where the SSB is estimated to be below 800,000 tonnes the Parties agree to set quotas for the directed fishery and for bycatches in other fisheries, reflecting a fishing mortality rate of less than 0.1 for 2 ringers and older and of less than 0.04 for $0-1$ ringers.
5. Where the rules in paragraphs 2 and 3 would lead to a TAC which deviates by more than $15 \%$ from the TAC of the preceding year the parties shall fix a TAC that is no more than $15 \%$ greater or $15 \%$ less than the TAC of the preceding year.
6. Notwithstanding paragraph 5 the Parties may, where considered appropriate, reduce the TAC by more than $15 \%$ compared to the TAC of the preceding year.
7. Bycatches of herring may only be landed in ports where adequate sampling schemes to effectively monitor the landings have been set up. All catches landed shall be deducted from the respective quotas set, and the fisheries shall be stopped immediately in the event that the quotas are exhausted.
8. The allocation of the TAC for the directed fishery for herring shall be $29 \%$ to Norway and $71 \%$ to the Community. The bycatch quota for herring shall be allocated to the Community.
9. A review of this arrangement shall take place no later than 31 December 2011.
10. This arrangement enters into force on 1 January 2009.

Also from January 2009 (EU Council Reg No 43/2009) high-grading and slipping of fish over the minimum landing size (as low as quota still exists) has been banned in EU waters. Discarding is illegal in Norwegian waters.

## H. 4 Sampling of commercial catch

Sampling of commercial catch is conducted by the national institutes. HAWG has recommended for years that sampling of commercial catches should be improved for most of the stocks. In January 2008, a new directive for the collection of fisheries data was implemented for all EU member states (Commission Regulations 2008/949/EC, 2008/199 and 2008/665). The provisions in the "data directive" define specific sampling levels. As most of the nations participating in the fisheries on herring assessed here have to obey this data directive, the definitions applicable for herring and the area covered by HAWG are given below:

| AREA | SAMPLING LEVEL PER 1000 T CATCH |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Baltic area (IIIa (S) and IIIb-c) | 1 sample of <br> which | $\mathbf{1 0 0}$ fish measured <br> and | $\mathbf{5 0}$ aged |
| Skagerrak (IIIa (N)) | $\mathbf{1}$ sample | $\mathbf{1 0 0}$ fish measured | 100 <br> aged |
| North Sea (IV and VId): | $\mathbf{1}$ sample | 50 fish measured | $\mathbf{2 5}$ aged |
| NE Atlantic and Western Channel ICES areas <br> II, V, VI, VII (excluding d) VIII, IX, X, XII, XIV | $\mathbf{1}$ sample | 50 fish measured | 25 aged |

Exemptions to the above mentioned sampling rules are:
Concerning lengths:
(1) the national programmeme of a Member State can exclude the estimation of the length distribution of the landings for stocks for which TACs and quotas have been defined under the following conditions:
(i) the relevant quotas must correspond to less than $5 \%$ of the Community share of the TAC or
to less than 100 tonnes on average during the previous three years;
(ii) the sum of all quotas of Member States whose allocation is less than $5 \%$, must account for
less than $15 \%$ of the Community share of the TAC.
If the condition set out in point (i) is fulfilled, but not the condition set out in point (ii), the relevant Member States may set up a coordinated programme to achieve for their overall landings the implementation of the sampling scheme described above, or another sampling scheme, leading to the same precision.

## Concerning ages:

(1) the national programmeme of a Member State can exclude the estimation of the age distribution of the landings for stocks for which TACs and quotas have been defined under the following conditions:
(i) the relevant quotas correspond to less than $10 \%$ of the Community share the TAC or to
less than 200 tonnes on average during the previous three years;
(ii) the sum of all quotas of Member States whose allocation is less than $10 \%$, accounts for less than $25 \%$ of the Community share of the TAC.

If the condition set out in point (i) is fulfilled, but not the condition set out in point (ii), the relevant Member States may set up a coordinated programmeme as men-
tioned for length sampling.
If appropriate, the national programmeme may be adjusted until 31 January of every year to take into account the exchange of quotas between Member States;

## H. 5 Terminology

The WG uses "rings" rather than "age" or "winter rings" throughout the report to denominate the age of herring, with the intention to avoid confusion. It should be observed that, for autumn spawning stocks, there is a difference of one year between "age" and "rings". HAWG in 1992 (ICES 1992/Assess:11) stated that
"The convention of defining herring age rings instead of years was introduced in various ICES working groups around 1970. The main argument to do so was the uncertainty about the racial identity of the herring in some areas. A herring with one winter ring is classified as 2-years-old if it is an autumn spawner, and one-year-old if it is a spring spawner. Recording the age of the herring in rings instead of in years allowed scientists to postpone the decision on year of birth until a later date when they might have obtained more information on the racial identity of the herring.

The use of winter rings in ICES working groups has introduced a certain amount of confusion and errors. In specifying the age of the herring, people always have to state explicitly whether they are talking about rings or years, and whether the herring are autumn- or spring spawners. These details tend to get lost in working group reports, which can make these reports confusing for outsiders, and even for herring experts themselves. As the age of all other fish species (and of herring in other parts of the world) is expressed in years, one could question the justification of treating WestEuropean herring in a special way. Especially with the present trend towards multispecies assessment and integration of ICES working groups, there might be a case for a uniform system of age definition throughout all ICES working groups.
However, the change from rings to years would create a number of practical problems. Data files in national laboratories and at ICES would have to be adapted, which would involve extra costs and manpower. People that had not been aware of the change might be confused when comparing new data with data from old working group reports. Finally, in some areas (notably Division IIIa), the distinction between spring- and autumn spawners is still hard to make, and scientists preferred to continue using rings instead of years.
The Working Group discussed at length the various consequences of a change from rings to years. The majority of the Group felt that the advantages of such a change did not outweigh the disadvantages, and it was decided to stick to the present system for the time being."
The text table below gives an example for the correlation between age, rings and year class for the different spawning types in late 2002:

| Year class (autumn spawners) | $2001 / 2002$ | $2000 / 2001$ | $1999 / 2000$ | $1998 / 1999$ |
| :--- | :--- | :--- | :--- | :--- |
| Rings | 0 | 1 | 2 | 3 |
| Age (autumn spawners) | 1 | 2 | 3 | 4 |
| Year class (spring spawners) | 2002 | 2001 | 2000 | 1999 |
| Rings | 0 | 1 | 2 | 3 |
| Age (spring spawners) | 0 | 1 | 2 | 3 |

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## Annex 4 - Stock Annex Western Baltic Spring Spawning Herring

Quality Handbook ANNEX: HAWG-herring WBSS
Stock specific documentation of standard assessment procedures used by ICES and relevant knowledge of the biology.

| Stock | Western Baltic Spring spawning herring (WBSS) |
| :--- | :--- |
| Working Group: | Herring Assessment Working Group for the Area <br> South of $62^{\circ} \mathrm{N}$ |
| Date: | 25.03 .2009 |
| Authors: | M. Cardinale, J. Dalskov, T. Gröhsler, H. Mosegaard, <br> M. van Deurs, J. Gröger |

## A. General

## A.1. Stock definition and biology

## Stocks

Herring caught in Division IIIa and the eastern North Sea is a mixture of two stocks: North Sea Autumn Spawners (NSAS) and Western Baltic Spring Spawners (WBSS). All spring-spawning herring in the eastern part of the North Sea (IVa and b east), Skagerrak (Subdivision 20), Kattegat (Subdivision 21) and the Western Baltic (Subdivisions 22,23 and 24) are treated as one stock, WBSS. The main spawning area of the WBSS is considered to be Greifswalter Bodden at Rügen (therefore also referred to as the Rügen-herring) (ICES, 1998), whereas NSAS utilizes spawning areas mainly along the British east coast (e.g. Burd, 1978; Zijlstra, 1969). The assessment also takes into account the few Norwegian Spring Spawners (NSS) caught in IVa north.

The contribution of Downs-herring to the mix-area of Division IIIa is likely to be relatively small (un-published data from otolith readings, DIFRES), and Downs-herring are therefore included under NSAS for the stock assessment of herring in Division IIIa and Subdivisions 22-24.

In the Western Baltic almost solely WBSS are being caught (few autumn spawners, however, have been observed). The majority of $2+$ ringers, however, migrate out of the area during quarter 2, to feed in Division IIIa and the North Sea and return in quarter 1 (Biester, 1979; Nielsen et al., 2001; van Deurs and Ramkaer, 2007).

In the Kattegat and the eastern Skagerrak, mainly 2+ ringers of the WBSS and 0 to 2ringers from the NSAS are being caught (ICES, 2004; ICES WD, 2006). The area provides a nursery habitat for juvenile NSAS (also other areas in the North Sea function as nursery areas), that assumable have drifted into the area as larvae (Burd, 1978; Heath et al, 1997). 0-1 ringer WBSS mainly uses nursery areas in Subdivision 22-24 and start to occur in the southern Kattegat as 1-ringers. The largest concentrations of herring during June/July seem to appear along the southern edge of the Norwegian Trench and in the area to the east of Læsø, in Kattegat (ICES, 2005; ICES, 2006). In $3^{\text {rd }}$ quarter large concentrations of $2+$ ringers of the WBSS are found in the southern Kat-
tegat and Subdivision 23 as they aggregate for the over-wintering, which mainly takes place in Subdivision 23 (Nielsen et al., 2001; Clausen et al., 2006).

In the eastern North Sea and the western Skagerrak mainly $2+$ ringers from WBSS and 1 to 2-ringer NSAS are being caught (Clausen et al., 2006). Peak catches of WBSS occur in quarter 3, during which the spawning stock of WBSS feed in these areas (ICES, 2002). According to the herring acoustic survey (ICES, 2006) the largest concentrations of herring in this area occur along the transition zone between the Skagerrak and the North Sea (ICES, 2006). Some $2+$ ringer NSAS are caught in $1^{\text {st }}$ and $4^{\text {th }}$ quarter, since part of the NSAS spawning stock over-winter in the Norwegian trench in this area. (Burd, 1978; Cushing and Bridger, 1966; Clausen et al., 2006).

In historic time several local late winter and spring spawning populations in the Skagerrak and the Kattegat has been described (e.g. Ackerfors, 1977; Rosenburg and Palmen, 1982). The largest of these seems to have reached extinction decades ago (ICES, 2004). Local spawning events during spring in a rather large number of fjords on the coast of Skagerrak and Kattegat, and both in Denmark, Sweden, and Norway are known still to occur regularly (HERGEN, EU project QLRT 200-01370, final report), but have been considered of minor importance for the herring fisheries (ICES, 2001). Recent genetic and morphological studies confirmed that these local spawning areas belong to distinct spawning populations (Bekkevold et al., 2005) and bear witness of a more complex composition of multiple populations than previously assumed. The migration behaviour of these populations is basically unknown and the methods for splitting them from the Rügen-herring in catches are still associated with large uncertainties (HERGEN, EU project QLRT 200-01370, final report). Also on the German coast of the Western Baltic we find more than the spawning grounds of Rügen. E.g. the spring spawning grounds of the Sleich Fjord (Kühlmorgen-Hille, 1983). It is unknown whether herring visiting spawning grounds in the Sleich Fjord belong to the Rügen-herring or should be considered an independent population. However, results presented by Biester (1979) and the population diversity found by Bekkevold et al. (2005) indicates that they too are likely to be genetically distinct from the Rügen-herring.

## Methods for stock separation

Experience within the Herring Assessment Working Group has shown that stock separation procedures based on size distributions often will fail.

The method for separating herring stocks in Norwegian samples, using vertebral counts (VC), as described in former reports of this Working Group (ICES 1991/ Assess:15), assumes that for NSAS, the mean vertebral count is 56.5 and for WBSS 55.8. The fractions of spring spawners (fsp) are estimated from the formula (56.50-v)/(56.555.8), where v is the mean vertebral count of the (mixed) sample with the restriction that the proportion should be one if $\mathrm{fsp}>=1$ and zero if $\mathrm{fsp}<=0$. The method is quite sensitive to within-stock variation (e.g. between year classes) in mean VC. The mean VC, of the previous mentioned local spring-spawners from the Norwegian Skagerrak fjords (it should be emphasised that this is not the Norwegian Spring Spawners alias Atlantic-Scandio Herring), is higher than for the NSAS (Rosenberg and Palmén, 1982; van Deurs, 2005), and will bias fsp estimates if present in the samples. The Norwegian samples used in the stock assessment are from the eastern North Sea. The local Norwegian spring spawners therefore only constitute a problem if they migrate to feeding areas in the eastern North Sea. Inconclusive results from a study of the tag pratsite A. simplex in herring, indicates that this may be the case (van Deurs and Ramkaer, 2007).

The introduction of otolith microstructure analysis in 1996-97 (Mosegaard and PoppMadsen, 1996) enables an accurate and precise split between three groups, autumn, winter and spring-spawners. Today this method is applied for the stock separation in all Danish and Swedish IIIa samples. However, different populations with similar spawning periods are not resolved with the present level of analysis. Different stock components that are not easily distinguished by their otolith microstructure (OM) are considered to have different mean vertebral counts (VC): E.g. the local Skagerrak winter/spring-spawners: 57 (Rosenberg and Palmén, 1982); Western Baltic Sea: 55.6 55.8 (Gröger and Gröhsler, 2001; ICES 1992/H:5). It should, however, be noted that the estimated stock specific mean VC varies somewhat among different studies, and the VC alone is not likely to be a successful tool for distinguishing between separate spring spawning populations in an assessment context .

Comparison between separation methods using frequency distributions of vertebral counts and otolith microstructure showed reasonable correspondence. Using this information the years from 1991 to 1996 was reworked in 2001, applying common splitting keys for all years by using a combination of the vertebral count and otolith microstructure methods (ICES, 2001). From 2001 and onwards, the otolith-based method only has been used for the Division IIIa.

Different methods of identifying herring stocks in the Division IIIa and Subdivisions 22-24 were recently evaluated in an EU CFP study project (EC study 98/026). The study involved several inter-calibration sessions between microstructure readers in the different laboratories involved with the WBSS herring. After the study was finished a close collaboration concerning reader interpretations has been kept between the Danish and Swedish laboratories. Sub-samples of the 2002 and 2003 Danish, Swedish, and German microstructure analyses were double-checked by the same Danish expert reader for consistency in interpretation. The overall impression is an increasingly good agreement among readers.
New molecular genetic approaches for stock separation are being developed within the EU-FP5 project HERGEN (EU project QLRT 200-01370, final report). Sampling of spawning aggregations during spring, autumn and winter has been carried out in 2002 and in 2003 in Division IIIa and in the Western Baltic at more than 10 different locations. The results point at a substantial genetic variation between North Sea and Western Baltic herring. As mentioned earlier, significant variation has also been found among spawning populations in Division IIIa and subdivision 22-24, which indicates the presence of multiple distinct spring spawning populations or subpopulations (Bekkevold et al., 2005). However, the substantial overlap in the genetic profiles of these sub-populations results in large uncertainties when attempting to estimate the proportional contribution of the individual spring spawning populations to the mix in Division IIIa.

For Subdivisions 22, 23 and 24 it is assumed that all individuals caught belong to the WBSS. However, after the introduction of OM analysis in 1996/97 it was discovered that in the western Baltic a small percentage of the herring landings might consist of autumn spawning individuals. Before molecular genetic methods became available for Atlantic herring the existence of varying proportions of autumn spawners in Subdivisions 22-24 in different years was considered a potential problem for the assessment, since they were thought to belong to the NSAS. Today the molecular genetic methods have revealed that they are more closely related to the WBSS than to the NSAS (HERGEN, EU project QLRT 200-01370, final report). Therefore, with the pre-
sent genetic perception in mind, when herring with OM indicating autumn hatch are found in subdivisions 22-24 these are treated as belonging to the WBSS stock.

OM analysis for stock splitting is a relatively time consuming method, furthermore, its potential for making splits, between the recently discovered complexity of different spring spawning populations, is very limited (un-published results, DIFFRES). Time has therefore been put into developing new, and more time efficient methods, for stock splitting. Under the EU-FP5 project HERGEN (EU project QLRT 200-01370, final report) a promising and time effective method based on otolith morphology are being developed. So far this work has showed that individual stocks and local populations display significantly different edge pattern of lobe formation in the otolith (the work was conducted on the saggitae otolith). The procedure involves photographing the shapes of the otolith edge and subsequent analysis in the photo treatment software Image Pro plus 5.0. However, so far the technique does not provide a way to efficiently split between spring spawning population in the mix-area of IIIa.

## A.2. Fishery

## Fleet definitions

The fleet definitions used since 1998 for the fishery in Division IIIa are:

- Fleet C: directed fishery for herring in which trawlers (with 32 mm minimum mesh size) and purse seiners participate.
- Fleet D: All fisheries in which trawlers (with mesh sizes less than 32 mm ) and small purse seiners, fishing for sprat along the Swedish coast and in the Swedish fjords, participate. For most of the landings taken by this fleet, herring is landed as by-catch.

Danish and Swedish by-catches of herring from the sprat fishery and the Norway pout and blue-whiting fisheries are listed under fleet D.

In SDs 22-24 most of the catches are taken in a directed fishery for herring and some as by-catch in a directed sprat fishery. All landings from SDs22-24 are treated as one fleet.

## Historical German fishing pattern

The overall German fishing pattern has changed in the last few years. Until 2000 the dominant part of German herring catches were caught in the passive fishery by gillnets and trapnets around the Rügen Island. Since 2001 the activities in the trawl fishery increased. Recently the landings by trawl reached a level of more than $50 \%$ of the total landings (2003: 63 \%, 2004: 52 \%, 2005: $57 \%$ and 2006: $64 \%$ ). The change in fishing pattern was caused by requirements for a fish factory on Rügen Island established in 2003 which can process 50000 t per year.

Investigation of new Danish fleet/metier description and the possibilities of improving the advice for the mixed stocks in IIIa (The IMHERSKA EU-project (Clausen et al., 2006))

An ecosystem approach to fisheries management should consider conservation of intra-specific variation due to population structure and life history variation. Knowledge of stock integrity is of unequivocal importance for sustainable fisheries management, since variable compositions in mixed areas together with asynchronous population dynamics may lead to over-fishing of individual stocks if not all components are managed to ensure (or achieve) sustainable exploitation.

A descriptive analysis of the Danish fleet dynamics during the last decade, in terms of the distribution of herring catches over fleets and at the overall activity of the vessels targeting herring in Division IIIa, together with an investigation of the fleet/metier specific exploitation of the individual stocks in Division IIIa was performed in the IMHERSKA EU project (Clausen et al., 2006).

For the descriptive analysis of the Danish fleet dynamics during the last decade, the fisheries identified in Ulrich and Andersen (2004) was modified accordingly, to get as much consistency with the previous HAWG work. Fisheries were identified using a 3-steps method using multivariate analysis of landings profile (target species) and trips descriptors (mesh size, season, and area). The data were based on logbook data and though considerable misreporting is suspected to take place between Division IIIa and the North Sea, the geographical patterns described below is believed to illustrate the fishery behaviour in general terms.

Figure A.2.1 illustrates the distribution of Danish herring landings in Division IIIa by vessel type and homeport (fleet) in 2004. From this 4 fleets were identified and Figure 3.1.2 shows the distribution of herring landings by fleet over selected years:

1 ) OTB_NSSK: trawlers from North Sea and Skagerrak harbours (Skagen included). This fleet is referred to as the Northern fleet.
2 ) PSB_NSSK: purse-seines from North Sea and Skagerrak harbours.
3 ) OTB_KAWB: trawlers from North Sjælland and Western Baltic (Subdivisions 22-24) harbours. This fleet is referred to as the Southern fleet.
4 ) OTH: all other vessels recorded for having caught herring in Division IIIa at least once a year. Given its low importance, this fleet is not kept further in the analysis.


Figure A.2.1 Danish landings in IIIa by vessel and homeport.

The spatial and temporal distribution of the two main stocks (NSAS and WBSS respectively) in the Subdivisions IVaE, IIIaN, IIIaS and Subdivisions 22-24 based on analysis of herring catch compositions from both commercial and scientific sampling in the period from 1999 to 2004 appear to be following certain patterns in terms of seasonality which in turn allow predictions of the mix of herring in the area. Furthermore, by using the above four fleets/metiers and disaggregating those further into industrial or commercial activities and looking at the stock composition in their catches within different seasons, stock selective metiers was identified (a stock selective metier was defined as: a metier with $80 \%$ or more of its landings constituting the same stock). Identifying such patterns, both in terms of the life-stage spatiality of WBSS and NSAS in division IIIa and adjacent areas, and in terms of fleets activity and inter-stock selectivity was a necessary prerequisite for any use of improved fleet- and stock-based management objectives. We have thus demonstrated that a more precise advice for the mixed stock in IIIa using elaborate fleet- and stock-based desegregations could be implemented. A projection method for predicting both stock- and me-tier-specific Fs is being developed accordingly.

## Historical Danish fishing pattern

The general dynamics of the Danish herring activities in Division IIIa can be summed up as the following points:

- During the first half of the 1990-ties, the activity was relatively local. The fleets were mostly fishing in their immediate waters. For some of the vessels mainly participating in the small meshed fisheries the fishery for herring for human consumption was a minor but stable activity.
- The second half of the 1990-ties was a period of extension. Both the Southern and Northern trawling fleets extended their activity to the Baltic, and decreased meanwhile their industrial activities in the Kattegat and Skagerrak, which induced reduced by-catches of herring. In the same period, the large purse seiners (most of the vessels are polyvalent) increased significantly their geographical mobility, with a majority of their effort being spent outside the traditional Danish fishing grounds in the North Sea and Division IIIa as they participated in fishery for blue whiting and Norwegian spring spawning herring.

The Swedish fleet definition is based on mesh size of the gear as for the Danish fleet. However, a recent change in the Swedish industrial fishery has occurred, as the Swedish industrial fishery has rapidly declined during the 1990's and it is currently no longer operating in the area. Therefore, there is no difference in age structure of the Swedish landings between vessel using different mesh sizes since both are basically targeting herring for human consumption.

## Changes in fishing technology and fishing patterns

Since 2001, the fishing pattern has changed in the German fleet. In former years, the main catch of herring was taken in the passive gears, bottom-set gillnets and trapnets. Recently the landings by trawl have reached a level of more than $50 \%$ of the total landings (2003: $63 \%, 2004: 52 \%, 2005: 57 \%$ and 2006: $64 \%$ ). This change is due to requirements from a new fish factory on the Rügen Island.

The Swedish industrial fishery rapidly declined during the 1990s and it is currently no longer operating in the area. Therefore, there is no difference in age structure of the landings between vessels using different mesh sizes since both are basically targeting herring for human consumption.

A descriptive analysis of the Danish fleet dynamics during the last decade, in terms of the distribution of herring catches over fleets and at the overall activity of the vessels targeting herring in IIIa, was performed in the IAMHERSKA (Improved Advice for the Mixed HERring stocks in the Skagerrak and Kattegat (ICES Division IIIa)) project (Ulrich-Rescan and Andersen 2006 WD 1 in ICES CM 2006/ACFM: 20). During the second half of the nineties, both the southern and northern trawling fleets extended their activity to the Baltic, and decreased meanwhile their industrial activities in the Kattegat and Skagerrak, which induced reduced by-catches of herring. In the same period, the large purse seiners (most of the vessels are polyvalent) increased significantly their geographical mobility, with a majority of their effort being spent outside the traditional Danish fishing grounds in the North Sea and Division IIIa as they participated in fisheries for blue whiting and Norwegian spring spawning herring.

The full consequence of the implementation of the ITQ system for herring is yet unknown as vessels still are changing status. However, a change in the behaviour in the Danish herring fishery indicates that vessels without an ITQ for herring are targeting a mixed sprat and herring fishery and land their catch for industrial purposes, whereas vessels with an ITQ for herring are primarily participating in the herring fishery for human consumption.

## A.3. Ecosystem aspects

Recent results from the HERGEN research-project on herring (HERGEN, EU project QLRT 200-01370, final report) reveals an increase in genetic distance between herring populations in the Baltic and successive populations in subdivisions 24, 22, 21, and 20 and finally the North Sea where genetic distance reach a maximum constant difference to the Baltic. Further, genetic differences are larger among populations within the Division IIIa and Western Baltic than among populations in the North Sea. The results also suggests that the herring spawning in spring on local spawning areas in the fjords of both the Western Baltic, the Kattegat, and the Skagerrak should be regarded as distinct spawning populations (or sub-populations) rather than as "strayers" from the Rügen-herring population. Furthermore, the contribution of these local spring spawning populations are considerable (Bekkevold et al., 2005; HERGEN, EU project QLRT 200-01370, final report).
By comparing five different Baltic herring stocks, temperature and SSB was shown as a the main predictors contributing to explain recruitment in the whole Baltic Sea, (Cardinale et al. 2009) except for Western Baltic herring where the Baltic Sea Index was the selected proxy in the final model. However, Baltic Sea Index is also known to be related to SST in the area.

## B. Data

## B.1. Commercial catch

Misreporting to fishing area still occurs. There is uncertainty about where the Danish landings for human consumption, reported from Division IIIa were actually taken. There is a high probability that these catches have been taken in the North Sea. Therefore, some of these catches have been transferred to the North Sea. Lastly, some landings reported as taken in the Triangle (Gilleleje, DK - Kullen, S - Helsingborg, S Helsingør, DK), may have been taken outside this area and listed under the Kattegat.

There is at present no information about the relevance of local herring stocks/populations in relation to the fisheries and their possible influence on the stock assessment. Recent evidence from genetic differentiation among spawning aggregations in the Skagerrak suggests a potential high representation of these local spawning stocks (Bekkevold et al., 2005). Other results suggest that at least the mature proportion of the different stock components to a large extent shares migration patterns and feeding areas (Ruzzante et al., 2006; van Deurs and Ramkaer, 2007).

## B.2. Biological parameters for assessment

Mean weights-at-age in the catch in the $1^{\text {st }}$ quarter were used as stock weights.
In order to check if this is a valid assumption and represents the actual weights in the stock, the index was compared to the average weights in the catch by age during the whole year. The relationship followed the expected pattern where the weight of the younger age classes in the catch are somewhat higher than in the stock as these are taken as an average over the whole year allowing for growth. From age-class 4 the relation between weight in catch and weight in stock followed a 1:1 line as expected. Thus the use of weight in the catch in quarter 1 is a sound indicator for the weight in the stock and does not give a biased representation of the stock.

The proportion of F and M before spawning was assumed constant between years. Fprop was set to be 0.1 and M-prop 0.25 for all age groups.

Natural mortality was assumed constant at 0.2 for all years and $2+$ ringers. A predation mortality of 0.1 and 0.2 was added to the 0 and 1 ringers, which resulted in an increase in their natural mortality to 0.3 and 0.5 , respectively (Table 3.6.4). The estimates of predation mortality were derived as a mean for the years 1977-1995 from the Baltic MSVPA (ICES 1997/J:2).

The maturity ogive was assumed constant between years:

| W-rings | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Maturity | $\mathbf{0 . 0 0}$ | $\mathbf{0 . 0 0}$ | $\mathbf{0 . 2 0}$ | $\mathbf{0 . 7 5}$ | $\mathbf{0 . 9 0}$ | $\mathbf{1 . 0 0}$ | $\mathbf{1 . 0 0}$ | $\mathbf{1 . 0 0}$ | $\mathbf{1 . 0 0}$ |

## B.3. Surveys

The summer Danish acoustic survey in Division IIIa is part of an annual survey covering the North Sea and Division IIIa in July-August. R/V DANA conducted the survey in Division IIIa. For each sub area the mean back scattering cross section was estimated for herring, sprat, gadoids and mackerel by the TS relationships given in the Manual for Herring Acoustic Surveys in ICES Division III, IV, and IVa (ICES 2002/G:02). Used in the final assessment.

The first joint acoustic survey was carried out with R/V 'Solea' in Subdivisions 22-24 in October 1987. Since 1989 the survey was repeated every year as a part of an international hydracoustic survey in the Baltic. The survey has been revised in 2007 and it now includes also SD 21. Used in the final assessment.

The IBTS $3^{\text {rd }}$ quarter survey in Division IIIa is part of the North Sea and Div. IIIa bottom trawl survey carried out in the $1^{\text {st }}$ and $3^{\text {rd }}$ quarter. The IBTS has been conducted annually in the $1^{\text {st }}$ quarter since 1977 and $3^{\text {rd }}$ quarters from 1991. From 1983 and onwards the survey was standardised according to the IBTS manual (ICES 2002/D:03). During the HAWG 2002 the IBTS survey data (both quarter) were revised from 1991 to 2002. Historical catch rates are heavily skewed and therefore the survey indices by winter rings 1-5 were calculated as geometric means from observed abundances ( $\mathrm{n} \cdot \mathrm{h}-$
${ }^{1}$ ) at age at trawl stations. However, inspections of the distributions of CPUE ( $\mathrm{n} \cdot \mathrm{h}^{-1}$ ) reveals that they are characterized by a relatively large number of low values, including true zeroes, but also occasional catches comprising large number of individuals. Statistical inference based on such data is likely to be inefficient or wrong unless an appropriate distribution is carefully chosen Generally, a quasi-Poisson distribution (with a log-link function in order to constraint the estimates of CPUE to be positive) and a so called zero inflated models (Minami et al. 2006; Martin et al., 2005) are used. While quasi-Poisson can treat zeroes and non-zeroes in the same models, zeroinflated models are expressed in two parts: the probability of being in a 'perfect-state' (e.g., no catch), and the probability of being in an 'imperfect-state' where positive events (e.g., catch) may occur (Minami et al. 2006). The perfect-state is usually modeled with a logistic, and a quasi-Poisson or a negative binomial distribution is assumed for the imperfect state. Those models are usually referred to as zero-inflated (ZIP and ZINB) models. Zero-inflated models are also attractive because they make a distinction between covariates associated with the perfect state (no catch) and covariates associated with the imperfect state in which catch can occur, but is not certain. Analysis is ongoing to test the use of ZIP and ZINB for estimating catch at age from IBTS dataset to be included in the next benchmark assessment. Thus, the IBTS indices were not used in the final assessment from 2008 and onwards. Not used in the final assessment.

The German herring larvae monitoring started in 1977 and takes place every year from March/April to June in the main spawning grounds of the spring spawning herring in the Western Baltic. These are the Greifswalder Bodden and adjacent waters. For the calculation of the number of larvae per station and area unit, the methods of Smith and Richardson (1977) and Klenz (1993) were used and projected to lengthclasses. Further details concerning the surveys and the treatment of the samples are given in Brielmann (1989), Müller and Klenz (1994) and Klenz (2002). Data revision was made in 2007 with a new method in calculating number at 20 mm . There was a high correlation between the indices N20 and HA_1 which are based on significantly different methods, areas and periods. Thus, results suggest that the index N20 is a suitable estimator of the new year-class of the spring spawning herring in ICES subdivision 22 - 24 (Oeberst et al, 2007, WD 7 in HAWG 2008 report). The time series now starts in 1992. Used in the final assessment.

## B.4. Commercial CPUE

## None

## B.5. Other relevant data

## None

## C. Historical Stock Development

Model used: ICA
Software used: FLICA

Model Options chosen:
No of years for separable constraint: 5
Reference age for separable constraint: 4
Constant selection pattern model: yes
$S$ to be fixed on last age: 1.0
First age for calculation of reference F: 3
Last age for calculation of reference F: 6
Relative weights-at-age: 0.1 for 0 -group, all others 1
Relative weights by year: all 1
Catchability model used: for all indices linear
Survey weighting: Manual all 1
Estimates of the extent to which errors in the age-structured indices are correlated across ages: all 1
No shrinkage applied
Input data types and characteristics:

| Type | Name | Year range | Age range | Variable from year to year Yes/No |
| :---: | :---: | :---: | :---: | :---: |
| Caton | Catch in tonnes | 1991- last data year | 0-8+ | Yes |
| Canum | Catch-at-age in numbers | 1991- last data year | 0-8+ | Yes |
| Weca | Weight-at-age in the commercial catch | 1991- last data year | 0-8+ | Yes |
| West | Weight-at-age of the spawning stock at spawning time. | 1991- last data year | 0-8+ | Yes, assumed as the Mw in the catch first quarter |
| Mprop | Proportion of natural mortality before spawning | 1991- last data year | 0-8+ | No, set to 0.25 for all ages in all years |
| Fprop | Proportion of fishing mortality before spawning | 1991- last data year | 0-8+ | No, set to 0.1 for all ages in all years |
| Matprop | Proportion mature at age | 1991- last data year | 0-8+ | No, constant for all years |
| Natmor | Natural mortality | 1991- last data year | 0-8+ | No, constant for all years |

Presently used Tuning data:

| Type | Name | Year range | Age range |
| :--- | :--- | :--- | :--- |
| Tuning fleet 1 | Danish Acoustic <br> Survey Div. IIIa | $\mathbf{1 9 9 0}$ - last year data | 3-6 |
| Tuning fleet 2 | German Acoustic <br> Survey SDs 22-24 | 1993 - last year data | 1-3 |
| Tuning fleet 3 | N20 larval survey, <br> Greifswalder Botten | 1992 - last year data | $\mathbf{0}$ |

## D. Short-Term Projection

Model used: Age structured
Software used: MFDP 1a
Initial stock size: ICA estimates of population numbers were used except for

- the numbers of 0-ringers in the last two years and the start year of the projection, where a geometric mean of the recruitment over the period of 5 years was taken
- the numbers of 1-ringers in the start of the projection, was taken as the projected 0-ringers in 2008

Natural mortality: The same values as in the assessment is used for all years
Maturity: The same values as in the assessment is used for all years
F and M before spawning: The same ogives as in the assessment is used for all years
Weight-at-age in the stock: Average weight of the three last years
Weight-at-age in the catch: Average weight of the three last years

Exploitation pattern: Average weight of the three last years not rescaled to the last year (Catch constraint)

Intermediate year assumptions: Status quo fishing mortality
Stock recruitment model used: None
Procedures used for splitting projected catches: Not relevant

## E. Medium-Term Projections

Model used: HCS
Software used: HCS
Initial stock size: ICA estimates of population numbers were used
Natural mortality: The same values as in the assessment is used for all years
Maturity: The same values as in the assessment is used for all years
F and M before spawning: The same values as in the assessment is used for all years
Weight-at-age in the stock: Average weight of the three last years
Weight-at-age in the catch: Average weight of the three last years

Exploitation pattern: Average weight of the three last years Intermediate year assumptions: Status quo fishing mortality

Stock recruitment model used: Hockey stick
Uncertainty models used:

1) Initial stock size:

2 ) Natural mortality:
3 ) Maturity:
4) F and M before spawning:

5 ) Weight-at-age in the stock:
6 ) Weight-at-age in the catch:
7 ) Exploitation pattern:
8 ) Intermediate year assumptions:
9 ) Stock recruitment model used:

## F. Long-Term Projections

Model used: none
Software used:
Maturity:
F and M before spawning:
Weight-at-age in the stock:
Weight-at-age in the catch:
Exploitation pattern:
Procedures used for splitting projected catches:

## G. Biological Reference Points

There are no precautionary approach reference points for this stock. Based on yield per recruit analysis and simulation carried out during HAWG (2007) and WKHMP (2008), a proxy for long term maximum sustainable exploitation rate should be a level of fishing mortality should not exceed $\mathrm{F}=0.25$.

Risk assessment performed in 2007

To address the issue of risk assessment with respect to simulation based optimizations carried out for IIIa herring in section 3.8 we implemented the following risk definition as given in the SGRAMA report of 2006 (ICES 2006/RMC:04) which is risk in a juridical sense:

Risk $=\mathrm{P}($ harmful event $) \times$ severity of harmful event

$$
\begin{equation*}
=\text { P(lower SSB limit undercut }) \times \text { EL } \tag{1}
\end{equation*}
$$

with expected loss (EL) being defined as

$$
\mathrm{EL}=\mathrm{E}\left[\mathrm{SSB}_{\text {lower limit }}-\mathrm{SSB}_{\text {estimated }} \mid \mathrm{SSB}_{\text {estimated }}<\mathrm{SSB}_{\text {lower limit }}\right] . \text { (2) }
$$

While this definition of risk is not only implemented as part of many national constitutions (for instance, of the German constitution; Schuldt 1997, Schulte 1999, Schulz et al. 2001) but is also commonly used in engineering, in natural or environmental sciences or in medicine (see, for instance, Burgmann 2004), in mathematical sciences however P (harmful event) is often solely used as a definition for risk. As we aim at specifying costs or loss from a political and economic perspective, Eq. (1) turns out to be the appropriate risk measure, as it contains a probability term specifying the chance or likelihood of a harmful event and a severity term quantifying the magnitude of the loss. Further information on the theory underlying risk assessment and risk management can be found in Burgmann (2004), Francis and Shotton (1997) and Lane and Stephenson (1997). For a formal treatment of quantitative risk assessment and management see McNeil (2005).

## H. Other Issues

## None

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# Annex 5 - Stock Annex Herring in the Celtic Sea and VIIj 

Quality Handbook Herring in Celtic Sea and VIIj
Stock specific documentation of standard assessment procedures used by ICES.

Stock: Herring in the Celtic Sea and VIIj
Working Group:
Herring Assessment Working Group for the area south of $62^{0}$
Date:
Authors: Afra Egan and Maurice Clarke

## A. General

The herring (Clupea harengus) to the south of Ireland in the Celtic Sea and in Division VIIj comprise both autumn and winter spawning components. For the purpose of stock assessment and management, these areas have been combined since 1982. The inclusion of VIIj was to deal with misreporting of catches from VIIg. The same fleet exploited these stocks and it was considered more realistic to assess and manage the two areas together. This decision was backed up by work by the ICES Herring Assessment Working Group (HAWG) in 1982 that showed similarities in age profiles between the two areas. In addition, larvae from the spawning grounds in the western part of the Celtic Sea were considered to be transported into VIIj (ICES, 1982). Also it was concluded that Bantry Bay in VIIj, was a nursery for fish of south coast (VIIg) origin (Molloy, 1968).

A study group examined stock boundaries in 1994 and recommended that the boundary line separating this stock from the herring stock of VIaS and VIIb be moved southwards from latitude $52^{\circ} 30^{\prime} \mathrm{N}$ to $52^{\circ} 00^{\prime} \mathrm{N}$ (ICES, 1994b). However, a recent study (Hatfield, et al 2007) examined the stock identity of this and other stocks around Ireland. It concluded that the Celtic Sea stock area should remain unchanged.

Some juveniles of this stock are present in the Irish Sea for their first year or two of life. Juveniles, which are believed to have originated in the Celtic Sea move to nursery areas in the Irish Sea before returning to the spawn in the Celtic Sea. This has been verified through herring tagging surveys, conducted in the early 1990s, (Molloy, et al 1993) and studies examining otolith microstructure (Brophy and Danilowicz, 2002). Age distribution of the stock suggests that recruitment in the Celtic Sea occurs first in the eastern area and follows a westward movement. After spawning herring move to the feeding grounds offshore (ICES, 1994). In VIIj herring congregate for spawning in autumn but little is known about where they reside in winter (ICES, 1994). A schematic representation of the movements and migrations is presented in Figure 1. Figure 2 shows the oceanographic conditions that will influence these migrations.

The management area for this stock comprises VIIaS, VIIg, VIIj, VIIk and VIIh. Catches in VIIk and VIIh have been negligible in recent years. The linkages between this stock and herring populations in VIIe and VIIf are unknown. The latter are man-
aged by a separate precautionary TAC. A small herring spawning component exists in VIIIa, though its linkage with the Celtic Sea herring stock area is also unknown.

## A.2. Fishery

## Historical fishery development

Coastal herring fisheries off the south coast of Ireland have been in existence since at least the seventeenth century (Burd and Bracken, 1965). These fisheries have been an important source of income for many coastal communities in Ireland. There have been considerable fluctuations in herring landings since the early 1900s.

In the Celtic Sea, historically, the main fishery was the early summer drift net fishery and the Smalls fishery which also took place in the summer. In 1933 several British vessels, mainly from Milford Haven, began to fish off the coast of Dunmore East and the winter fishery gained importance. The occurrence of the world war changed the pattern of the herring fishery further with little effort spent exploiting herring in the immediate post war years (Burd and Bracken, 1965). Landings of herring off the south west coast increased during the 1950s.

In 1956 Dunmore East was considered as the top herring port in Ireland with over 3,000 t landed. This herring was mainly sold to the UK or cured and sent to the Netherlands (Molloy, 2006). During this time many boats from other European countries began to exploit herring in this area during the spawning period. This continued until the 1960s when catches began to fall. In 1961 the Irish fishery limits changed whereby non-Irish vessels were prohibited from fishing in the inshore spawning grounds (Molloy, 1980). Consequently, continental fleets could no longer exploit herring on the Irish spawning grounds. They had to purchase herring from Irish vessels in order to meet requirements (Molloy, 2006).

During the period from 1950-1968 the fleet exploiting the stock changed from mainly drift and ring nets to trawls. Further fluctuations in the landings were evident during this time with high quantities of herring landed from 1966-1971 (Molloy, 1972). In the mid-sixties, the introduction of mid-water pair trawling led to greater efficiency in catching herring and this method is still employed today. Overall the 1960s saw a rise in herring landings with 1969 seeing a rise to 48,000 t. The North Sea herring fisheries were becoming depleted and several countries were turning to Ireland to supply their markets. Prices also increased and additional vessels entered the fleet (Molloy, 1995). Increases in effort led to increased catches initially but this did not continue and the decline of the fishery began.

## Fishery in recent years

In the past, fleets from the UK, Belgium, The Netherlands and Germany as well as Ireland exploited Celtic Sea herring. In recent years however this fishery has been prosecuted entirely by Ireland. This fishery is managed by the Irish "Celtic Sea Herring Management Advisory Committee", established in 2000 and constituted in law in 2005.

The Irish quota is managed by allocating individual quotas to vessels on a weekly basis. Participation in the fishery is restricted to licensed vessels and these licensing requirements have been changed. Previously, vessels had to participate in the fishery each year to maintain their licence. Since 2004 this requirement has been lifted. This has been one of the contributing factors to the reduction in number of vessels participating in the fishery in recent seasons (ICES, 2005b). Fishing is restricted to the period Monday to Friday each week, and vessels must apply a week in advance before
they are allowed to fish in the following week. Triennial spawning box closures are enshrined in EU legislation (Figure 3).

The stock is exploited by two types of vessels, larger boats with RSW storage and smaller dry hold vessels. The smaller vessels are confined to the spawning grounds (VIIaS and VIIg) during the winter period. The refrigerated seawater (RSW) tank vessels target the stock inshore in winter and offshore during the summer feeding phase (VIIg). There has been less fishing in VIIj in recent seasons.

The fleet can be classified into four categories of vessels:
Category 1: "Pelagic Segment". Refrigerated seawater trawlers

Category 2: "Polyvalent RSW Segment".

Category 3: "Polyvalent Segment".

Category 4: Drift netters.

Refrigerated seawater or slush ice trawlers

Varying number of dry hold pair trawlers,

A negligible component in recent years, very small vessels

The term "Polyvalent" refers to a segment of the Irish fleet, entitled to fish for any species to catch a variety of species, under Irish law. Since 2002 fishing has taken place in quarter 3, targeting fish during the feeding phase on the offshore grounds around the Kinsale Gas Fields. These fish tend to be fatter and in better condition than winter-caught fish. In 2003 the fishery opened in July on the Labadie Bank and caught large fish. In 2004-2006 it opened in August and in 2007 and 2008 in September. Only RSW and bulk storage vessels can prosecute this fishery. Traditional dryhold boats are unable to participate.

In recent years, the targeting fleet has changed. The fleet size has reduced but an increasing proportion of the catch is taken by RSW and bulk storage vessels and less by dry-hold vessels. There has been considerable efficiency creep in the fishery since the 1980s with greater ability to locate fish.

## A.3. Ecosystem aspects

The ecosystem of the Celtic Sea is described in ICES WGRED (2007). The main hydrographic features of this area as they pertain to herring are presented in Figure 2.

Temperatures in this area have been increasing over the last number of decades. There are indications that salinity is also increasing (ICES, 2006). Herring are found to be more abundant when the water is cooler while pilchards favour warmer water and tend to extend further east under these conditions (Pinnegar, et al 2002).

However, studies have been unable to demonstrate that changes in the environmental regime in the Celtic Sea have had any effect on productivity of this stock.

Herring larval drift occurs between the Celtic Sea and the Irish Sea. The larvae remain for a period as juveniles before returning to the Celtic Sea. Catches of herring in the Irish Sea may therefore impact on recruitment into the Celtic Sea stock (Molloy, 1989). Distinct patterns were evident in the microstructure and it is thought that this is caused by environmental variations. Variations in growth rates between the two areas were found with Celtic Sea fish displaying fastest growth in the first year of life. These variations in growth rates between nursery areas are likely to impact recruitment (Brophy and Danilowicz, 2002). Larval dispersal can further influence maturity at age. In the Celtic Sea faster growing individuals mature in their second year (1 w.
ring) while slower growing individuals spawn for the first time in their third year (2 winter ring). The dispersal into the Irish Sea which occurs before recruitment and subsequent decrease in growth rates could thus determine whether juveniles are recruited to the adult population in the second or third year (Brophy and Danilowicz, 2003).

The spawning grounds for herring in the Celtic Sea are well known and are located inshore close to the coast. These spawning grounds may contain one or more spawning beds on which herring deposit their eggs. Individual spawning beds within the spawning grounds have been mapped and consist of either gravel or flat stone (Breslin, 1998). Spawning grounds tend to be vulnerable to anthropogenic influences such as dredging and sand and gravel extraction. The main spawning grounds are displayed in Figure 4, whilst the distributions of spawning and non-spawning fish are presented in Figure 5.

## By Catch

By catch is defined as the incidental catch of non target species. There are few documented reports of by catch in the Celtic Sea herring fishery. A European study was undertaken to quantify incidental catches of marine mammals from a number of fisheries including the Celtic Sea herring fishery. Small quantities of non target whitefish species were caught in the nets. Of the non target species caught whiting was most frequent ( $84 \%$ of tows) followed by mackerel ( $32 \%$ ) and cod ( $30 \%$ ). The only marine mammals recorded were grey seals (Halichoerus grypus). The seals were observed on a number of occasions feeding on herring when the net was being hauled and during towing. They appear to be able to avoid becoming entangled in the nets. It was considered unlikely by Berrow, et al 1998, that this rate of incidental catch in the Celtic Sea would cause any decline in the Irish grey seal population. Results from this project also suggested that there was little interaction between the fishing vessels and the cetaceans in this area. Occasional entanglement may occur but overall incidental catches of cetaceans are thought to be minimal (Berrow, et al 1998). The absence of any other by caught mammals does not imply that by catch is not a problem only that it did not occur during this study period (Morizur, et al 1999).

## Discards

Catch is divided into landings (retained catch) and discards (rejected catch). Discards are the portion of the catch returned to the sea as a result of economic, legal, or personal considerations (Alverson et al 1994). In the 1980s a roe (ovary) market developed in Japan and the Irish fishery became dependent on this market. This market required a specific type of herring whose ovaries were just at the point of spawning. A process developed whereby large quantities of herring were slipped at sea. This type of discarding usually took place in the early stages of spawning and was reduced by the introduction of experimental fishing (Molloy, 1995). This market peaked in 1997 and has been in decline since with no roe exported in recent years. Markets have changed with the majority of herring going to the European fillet market.

Presently there are no estimates of discards for this fishery used in assessments. Berrow, et al 1998 also looked at the issue of discarding during the study on by catch. The discard rate was found to be $4.7 \%$ and this compares favourably with other trawl fisheries. Possible reasons for discarding were thought to be the market requirements for high roe content and high proportions of small herring in the catch. Overall this study indicated that the Celtic Sea herring fishery is very selective and that discard rates are well within the figures estimated for fishery models.

Since the demise of the roe fishery, it is considered that the incentive to discard is less. However it is known that discarding still takes place, in response to a constrained market situation.

## B. Data

## B.1. Commercial Catch

The commercial catch data are provided by national laboratories belonging to the nations that have quota/fisheries for this stock. In recent years, only Ireland has been catching herring in this area, and the data are derived entirely from Irish logbook data. Figure 6 shows the trends in catches over the time series. Ireland acts as stock coordinator for this stock. Commercial catch at age data are submitted in Exchange sheet v 1.6.4. These data are processed either using SALLOCL (Patterson, 1998b), or using ad hoc spreadsheets, usually the latter. The relevant files are placed on the ICES archive each year

## Intercatch

Since 2007, InterCatch, which is a web-based system for handling fish stock assessment data, was also used. National fish stock catches are imported into InterCatch. Stock coordinators then allocate sampled catches to unsampled catches, aggregate them to stock level and download the output. The InterCatch stock output can then be used as input for the assessment models. The comparisons to date have been very good and it is envisaged that this system will replace SALLOCL and other previously used systems.

## B. 2 Biological

## Sampling Protocol

Sampling is performed as part of commitments under the EU Council Regulation 1639/2001.Sampling (of the Irish catches) is conducted using the following protocol

- Collect a sample from each pair of boats that lands. Depending on the size range a half to a full fish box (depending on size range) is sufficient. If collecting from processor make sure sample is ungraded and random.
- Record the boat name, ICES area, fishing ground, date landed for each sample. If possible find out roughly how much the boat landed.
- Randomly take 75 fish for ageing. Record length in 0.5 cm , weight, sex, maturity (use maturity scale for guideline). Extract otolith taking care not to break tip, store it in otolith tray. Make sure the tray is clean and dry.
- Record a tally for the 75 aged fish under "Aged Tally" on datasheet.
- Measure the remaining fish and record a tally on the measured component of datasheet


## Ageing Protocol

Celtic Sea herring otoliths are read using a stereoscopic microscope, using reflected light. The minimum level of magnification ( $15 x$ ) is used initially and is then increased to resolve the features of the otolith. Herring otoliths are read within the range of $20 x$ $-25 x$. The pattern of opaque (summer) and translucent (winter) zones is viewed. The winter (translucent) ring at the otolith edge is counted only in otoliths from fish
caught after the $1^{\text {st }}$ April. This "birth date" is used because the assessment year for Celtic Sea and Division VIIj herring runs from this date to the 31 st March of the following year (ICES, 2007). This ageing and assessment procedure is unique in ICES. A fish of 2 winter rings is a 3 year old. This naming convention applies to all ICES herring stocks where autumn spawning is a significant feature.

## Age composition in the catch

In recent years there is a decreasing proportion of older fish present in the catch. Figure 7 shows the age composition of the catches over the time series. It is clear that there is a truncation of older age classes in recent years.

## Precision in Ageing

Precision estimates from the ageing data were carried out in the HAWG in 2007, for the 2006/2007 season (ICES, 2007). Results found that CVs are highest on youngest and oldest ages that are poorly represented in the fishery. The main ages present in the fishery had low CVs, of between $5 \%$ and $13 \%$, which is considered a very good level of precision. In the third and the fourth quarter, estimates of 1 wr on CS herring were also remarkably precise. An overall precision level of 5\% was reached in Q1 and Q4 in the 2007/2008 season.

## Mean Weights and Natural Mortality

An extensive data set on landings is available from 1958. Mean weights at age in the catch in the 4 th and 1st quarter are used as stock weights. Trends in mean weights at age in the catches are presented in Figure 8, and for weights in the spawning stock in Figure 9. Clearly there has been a decline in mean weights since the early 1980s, to the lowest values observed.

The natural mortality is based on the results of the MSVPA for North Sea herring. Natural mortality is assumed to be as follows:

| 1 ringer | 1 |
| :--- | :--- |
| 2 ringer | 0.3 |
| 3 ringer | 0.2 |
| 4 and subsequent ringer | 0.1 |

## Maturity Ogive

Clupea harengus is a determinate one-batch spawner. In this stock, the assessment considers that $50 \%$ of 1 ringers are mature and $100 \%$ of two ringers mature. The maturity ogive calculated from acoustic survey data in 2007 shows that $58 \%$ of 1 ringers are mature and $99 \%$ of 2 ringers. Lynch (in prep) has also shown that more than $50 \%$ of 1 ringers are mature in recent years. It is to be noted that the fish that recruit to the fishery as 1-ringers are probably precocious early maturing fish. Late maturing 1ringers may not be recruited. Thus maturity at 1-ringer in the population as a whole may be different to that observed in the fishery. Late maturing 1-, 2- and even 3ringers may recruit from the Irish Sea.

## B.3. Surveys

## Acoustic

Acoustic surveys have been carried out on this stock from 1990-1996, and again from 1998-2008. During the first period, two surveys were carried out each year designed
to estimate the size of the autumn and winter spawning components. The series was interrupted in 1997 due to the non-availability of a survey vessel. Since 2005, a uniform design, randomised survey track, uniform timing and the same research vessel have been employed. A summary of the acoustic surveys is presented in Table 1.

## Revision of acoustic time series

A review of the acoustic survey programme was conducted to check the internal consistency of the previous surveys and produce a new refined series for tuning the assessment (Doonan, unpublished). The old survey abundance at age series is presented in Table 2 and the revised survey time series is shown in the Table 3 (ICES, 2006).

The surveys were divided into two series, early and late, based on how far from the south coast of Ireland the transects extended. The early group, 1990-91 to 1994-95, extended to about 15 nautical miles offshore with two surveys, one in autumn and another in winter. This design aimed to survey spawning fish close inshore with two surveys, the results of which could be added, the two legs covering the two main spawning seasons. The off shore limits were extended in 1995 and some of these surveys had more fish off shore than close inshore. This changed the catchability, suggesting the later series should be separated from the earlier one. Consequently the years before 1995 were removed. This is not considered to be a problem because the earlier series would contribute little to the assessment anyway.

The autumn surveys did not cover the southwest Irish coast of VIIj in all years (3 years missing). In order to correct for this, the missing values were substituted with the mean of the available western bays SSB estimates, 7800 t ( 11 values, range from 0 to 16000 t ). Numbers-at-age in these surveys were adjusted upwards by the ratio of the adjusted SSB in the SW to the south coast SSB. The current time series included autumn surveys only.

Analysis errors were found in the surveys from 1998 onwards. The 2003 biomass (SSB, 85500 t ) was re-analysed after the discovery of errors in the spreadsheets used to estimate biomass. The errors affected the calculation of the weighted mean of the integrated backscatter when positive samples had lengths shorter than the base one (here, 15 minutes) and the partitioning of the backscatter for a mixture of species. Also, no account was taken of different sampling frequencies within a $10 \times 20$ minute cell (the analysis unit). The 2003 SSB came mainly from two cells that included an intensive survey in Waterford Harbour and these cells had an SSB of about 68000 t , which was reduced to 7300 t when all errors were corrected. There were some minor corrections in three other cells. The revised total biomass was 24000 t and the revised spawning biomass was 22700 t .

In addition, the cell means took no account of the implicit sampling area of transects so that the biomass coming from a large sample value depended on the number of transects passing through the cell. The data were re-analysed using mean herring density by transect as the sample unit and dividing the area into strata based on transect spacing. Areas with no positive samples were excluded from the analysis (since they have zero estimates). Zigzags in bays were analysed as before. For each stratum, a mean density was obtained from the transect data (weighted by transect length) and this was multiplied by the stratum area to obtain a biomass and numbers-at-age. The overall total was the sum of the strata estimates. The same haul assignments as in the original analysis were used. At the same time, a CV was obtained based on
transect mean densities, i.e. a survey sample error. For surveys before 1998 and the western part survey in 2002, a CV was estimated using;

$$
\sqrt{\log \left(1.3^{2}\right) / n}
$$

where n is the number of positive sample values ( 15 minute of survey track) from Definite and Probably Herring categories. This was based on the data from the autumn surveys in 1998, 2000, 2001, 2002, and 2005.

## Current acoustic survey implementation

The acoustic data are collected using the Simrad ER60 scientific echosounder. The Simrad ES-38B ( 38 KHz ) split-beam transducer is mounted within the vessels drop keel or in the case of a commercial vessel mounted within a towed body. The survey area is selected to cover area VIIj, and the Celtic Sea (areas VIIg and VIIaS). Transect spacing in these surveys has varied between 1 to 4 nmi . For bays and inlets in the southwest region (VIIj) a combined zigzag and parallel transect approach was used to best optimise coverage. Offshore transect extension reached a maximum of 12 nmi , with further extension where necessary to contain fish echotraces within the survey area.

The data collected is scrutinised using Echoview ${ }^{\circledR}$ post processing software. The allocated echo integrator counts ( Sa values) from these categories were used to estimate the herring numbers according to the method of Dalen and Nakken (1983). The following target strength to fish length relationships is used for herring.

$$
\mathrm{TS}=20 \log \mathrm{~L}-71.2 \mathrm{~dB} \text { per individual }(\mathrm{L}=\text { length in } \mathrm{cm})
$$

## Acoustic Survey Time Series

The acoustic survey design has been standardised and the timing has been consistent each year since 2005. The 2002 and 2003 surveys had similar timing and are comparable to the uniform time series. In the benchmark assessment (2007) the time series used was from 1995-2006. At the time of the benchmark, there were not enough comparable consistent surveys available for tuning. In 2009, four consistent surveys (2005-2009) and two additional fairly consistent surveys (2002-2003) were available. This shorter series from 2002-2008 is considered the most consistent available.

## Irish Groundfish Survey

The IGFS is part of the western IBTS survey and has been carried out on the $R V$ Celtic Explorer since 2003. The utility of the IGFS as a tuning series was investigated (Johnston and Clarke, 2005 WD). Strong year effects were evident in the data. Herring were either caught in large aggregations or not at all. The signals from this survey were very noisy, but when a longer time series is developed, it will at least provide qualitative information. The absence of the 2001 year class was supported in the survey data in 2004.

## French EVHOE Survey

The Herring Assessment Working group in 2006 had access to data from the French EVHOE quarter 4 western IBTS survey (GOV trawl). The French survey series is from 1997 to 2005 and displayed very variable observed numbers at age between years. Consequently, further exploration of the series was not performed.

## UK Quarter 1 survey

The UK quarter 1 survey was also explored and strong year and age effects, particularly at 2- and 5-ringers were found. Due to strong year and age effects and because it was discontinued in 2002 this survey is considered unsuitable as a recruit index (ICES 2006:ACFM 20).

While these data are useful for comparisons between surveys, as with the Irish data, at the moment it is difficult to see how these data can be used in an assessment. The data, particularly towards the end of the time series are very noisy and the absence of very small (juvenile) fish, particularly 1 ringers for the majority of time series is not encouraging (Johnston and Clarke, 2005).

## Irish and Dutch juvenile herring trawl surveys

Juvenile herring surveys were carried out from 1972 - 1974 by Dutch and Irish scientists. These surveys aimed to get information on the location and distribution of young herring. They were also used to examine if young herring surveys in the Irish Sea could provide abundance indices for either the Irish Sea or Celtic Sea stocks. Further young fish surveys were carried out in the Irish Sea from 1979 - 1988. They were discontinued when it was decided that it was not possible to use the information as recruitment indices for the Celtic Sea or Irish Sea stocks despite earlier beliefs (Molloy, 2006). This was because it was not known what proportion of the catches should be assigned to each stock.

## Northern Ireland GFS surveys

These surveys take place in quarters 1 and 3 each year. Armstrong et al (2004) presented a review of these surveys. They are likely to be useful if the natal origin can be established. Further work in this area is required to examine if this survey can be used as a recruit index for Celtic Sea Herring.

## Larval Surveys

Herring larval surveys were conducted in the Celtic Sea between October and February from 1978 to 1985 with one further survey carried out in 1989. These surveys provided information on the timing of spawning and on the location of the main spawning events as well as on the size of autumn and winter spawning components of the stock. The larval surveys carried out after the fishery reopened in 1982 showed an increase in the spawning stock (Molloy, 1995).

The surveys covered the south coast and stations were positioned 8 nautical miles apart in a grid formation. A Gulf III sampler, with $275 \mu \mathrm{~m}$ mesh was used to collect the samples. The total abundance of $<10 \mathrm{~mm}$ larvae (prior to December $15^{\text {th }}$ ) or $<11 \mathrm{~mm}$ (after December $15^{\text {th }}$ ) was calculated by raising the numbers per $\mathrm{m}^{2}$ by the area represented by each station. The mean abundance of $<11 \mathrm{~mm}$ larvae in December - February gave the winter index which when multiplied by 1.465 and added to the Autumn index to give a single index of the whole series (Grainger et al 1982). Larval surveys have not been undertaken in this area since 1989 and until the acoustic survey became established, no survey was available to tune the assessment.

## B.4. Commercial CPUE

In the 1960s and 1970s CPUE (Catch per unit effort) data from commercial herring vessels were used as indices of stock abundance because there were no survey data available. These data provided an index of changes that were occurring in the fishery at the time. CPUE data were used to tune the assessment (Molloy, 2006). However it is likely that the decline in the stock in the 1970s was not picked up in the CPUE until it was at an advanced stage. It is now demonstrated that CPUE data does not provide an accurate index of herring abundance, as they are a shoaling fish.

## C. Historical Stock Development

## Time Periods in the Fishery

This fishery can be divided into time periods. A number of factors have changed in this fishery overtime such as the markets, discards and the water allowance. These changes have implications for the trustworthiness of the catch data used in the assessment. The time periods are presented in the Table 4. The recent biological history of the stock is presented in Table 5. It is clear that growth rate has changed over time. Mean length and mean weight at age have declined by about $15 \%$ and $30 \%$ respectively since the late 1970s. Fish are shorter and lighter at age now than at any time in the series. Trends in mean weights in the catch and in the stock are presented in Figure 8 and Figure 9.

## Exploration of basic data

Data exploration consisted of examining a number of features of the basic data. These analyses included log catch ratios, cohort catch curves in survey and catch at age series. Log catch ratios were constructed for the time series of catch at age data, as follows:

$$
\log [C(a, y) / C(a+1, y+1)]
$$

These are presented in Figure 10. It can be seen that 1-ringers, and the oldest ages, have a noisy signal, being poorly represented in the catches. There was an increase in ratios in 1998, that seems quite abrupt. Overall there is a trend towards greater mortality in recent years. The increased mortality visible in the older ages corresponds with the truncation in oldest ages in the catch at age profile. It can also be seen that the gross mortality signal was low in 2002, corresponding to the big decrease in catch in that year. The signal increased again in 2003, concomitant with increasing catch.
Cohort catch curves across all ages were constructed using the catch at age data and are presented in Figure 11. The total mortality $(Z)$ over ages 2-7 for the cohorts 19581997 is presented in Figure 12 and in Table 6. Fluctuations are evident with an increasing trend in recent years. Total mortality was low for cohorts 1956 to 1964. Cohorts in the late 1960s seem to display higher Z, but those from 1975 to 1982 displayed the highest $Z$ ( 0.6 to 1.1). The most recent year classes for which enough observations are available (1991-1997) show higher Z again, in the range about 0.6 to 1.0. Cohort catch curves were also constructed from the catch at age data across ages 2-5 (Figure 13) and the survey data for year classes where enough data were available (Figure 14). A secondary peak corresponding to the 2003/2004 season is obvious in the cohort catch curves. The same patterns in raw mortality are visible, but the Zs from the acoustic survey are somewhat higher than those from the commercial data. This may be explained as differing catchability between the two, and it should be noted when interpreting the assessment results below.

In conclusion only the cohorts from before the stock collapsed and a few from the late 1980s contributed many of the older fish that appear in the catches. Raw mortality signals, from cohort catch curves suggest that some of the recent year classes have displayed a higher total mortality.

Assessments 2007, 2008 and 2009
In 2007, a benchmark assessment used a variety of models including ICA (Patterson, 1998), separable VPA, XSA, CSA and Bayesian catch at age methods. In addition an analysis of long term dynamics of recruitment was conducted. Simulations of various fishing mortalities were conducted based on stock productivity. Though no final model formulation was settled upon, the assessment provided information on trends. ICA was preferred to XSA because it is more influenced by younger ages that dominate the stock and fishery, and because of consistency. The settings that had been used before 2007 were found to produce the most reasonable diagnostics.

In 2007 it was considered that the assumption that a constant separable pattern could be used may not have been valid and it was recommended that future benchmark work should consider models that allow for changes in selection pattern.

Also in 2007 a reduction of the plus group to $7+$ was recommended. This change did not achieve better diagnostics in 2007, but exploratory assessments in 2008 did find that this change improved the diagnostics.

In 2008 and 2009, the working group continued to explore different assessment settings in ICA. The working group treated these explorations as extensions of the benchmark of 2007. In 2008 ICA was replaced by FLICA and the same stock trajectories were found in each.

In 2009 a final analytical assessment was proposed and was conducted using FLICA(flr-project.org). This assessment was based on exploratory work done in 2008 and 2009. The refinements to the benchmark assessment of 2007 were as follows:

- Further reduction of plus group to 6+
- Exclusion of acoustic surveys before 2002, because a sufficient series of comparable surveys was now available.

The assessment showed improved precision and coherence between the catch at age and the survey data. The survey residuals were lower since 2002 which is reflected in better tuning diagnostics. The stock trajectories, based on this assessment are presented in Figure 15.

The model formulation used for ICA in the 2007 benchmark and the final assessment carried out in 2009 are presented in the table below.

| ICA Settings | 2007 Benchmark | Final Assessment 2009 |
| :--- | :--- | :--- |
| Separable period | 6 years (weighting $=1.0$ for <br> each year) | 6 years (weighting $=1.0$ for <br> each year) |
| Reference ages for separable <br> constraint | 3 | 3 |
| Selectivity on oldest age | 1.0 | 1.0 |
| First age for calculation of mean F | 2 | 2 |
| Last age for calculation of mean F | 6 | 5 |
| Weighting on 1 ringers | 0.1 | 0.1 |


| Weighting on other age classes | 1.0 | 1.0 |
| :--- | :--- | :--- |
| Ages for acoustic abundance <br> estimates | $2-5$ | $2-5$ |
| Plus group | 7 | 6 |

## Estimation of terminal year SSB

In this stock the procedure for calculation of terminal year SSB is different to other stocks. Recruits (1-ring) are poorly represented in the catch and only one observation of their abundance is available. Yet $50 \%$ of these are considered mature and they make an important contribution to the SSB. Therefore an adjustment is made, by replacing 1-ring abundance from ICA.out with GM recruitment. Examination of recent patterns shows that recruitment has fluctuated around the average 1995-2006. Therefore the GM recruitment estimate 1995-2006 is used. SSB is recalculated using GM estimate*stock weight at 1-ring* ${ }^{*}$ maturity at 1-ring.

Input data types and characteristics:

| TYPE | NAME | YEAR <br> RANGE | AGE <br> RANGE | VARIABLE FROM YEAR <br> TO YEAR <br> YES/NO |
| :--- | :--- | :--- | :--- | :--- |
| Caton | Catch in tonnes | $1958-$ <br> 2008 | $1-6+$ | Yes |
| Canum | Catch at age in numbers | $1958-$ <br> 2008 | $1-6+$ | Yes |
| Weca | Weight at age in the <br> commercial catch | $1958-$ <br> 2008 | $1-6+$ | Yes |
| West * | Weight at age of the spawning <br> stock at spawning time. | $1958-$ <br> 2008 | $1-6+$ | Yes |
| Mprop | Proportion of natural <br> mortality before spawning | $1958-$ <br> 2008 | $1-6+$ | No |
| Fprop | Proportion of fishing <br> mortality before spawning | $1958-$ <br> 2008 | $1-6+$ | No |
| Matprop | Proportion mature at age | $1958-$ <br> 2008 | $1-6+$ | No |
| Natmor | Natural mortality | $1958-$ <br> 2008 | $1-6+$ | No |

* mean weights in the stock in the new plus group were re-weighted using catch numbers at age.


## Tuning data:

| TYPE | NAME | YEAR RANGE | AGE RANGE |
| :--- | :--- | :--- | :--- |
| Acoustic Survey | CSHAS | $2002-2008$ | $2-5$ |

## Analysis of productivity over time

To account for the influence of the ecosystem on the productivity of this herring stocks (ICES, 2007, Chapter 1) the methods of Nash and Dickey-Collas (2005) were applied. The recruit per spawner ratio was calculated. These calculations formed the basis for the detection of periods of high and low production of the stock (Figure 17).

The next step was to calculate the net and surplus production of the whole stock, including the recruits and the growth of all non-recruits, the natural and the fishing mortality. To subtract the influence of the spawning stock biomass a hockey stick and a Ricker stock recruitment relationship were fitted to the data to obtain the residuals of the recruits of a given year. The residuals were used to remove the year effect from the estimation of the stock size and to gain the net production and the surplus production respectively without the effect of the SSB on the number of recruits. Contrary to ICES (2007, Technical Minutes) the stock recruit model is not presented. This is because the model is not considered a good fit to the data and because the aim of this analysis is to examine recruitment, having removed the effect of SSB.

The data used in this analysis was derived from the assessment outputs from the HAWG in 2006 (ICES HAWG, 2006, Table 1.8.3.1).

Calculation of the surplus production

$$
\mathrm{Ps}=\mathrm{Br}+\mathrm{Bg}-\mathrm{M}
$$

where Br is the biomass of the recruits, Bg the gain of biomass due to growth of all fish excluding the recruits and $M$ the natural mortality. The net production equals the surplus production minus the fishing mortality (F).

The Celtic Sea herring stock had a low productivity throughout the whole time series, compared to other stocks (ICES, 2007). The net and surplus production is very noisy displaying neither clear trend. The impact of a varying F was tested using the Hockey Stick stock recruitment relationship. The stock showed variable production over time (Figure 17). It can be seen that $\mathrm{F}_{0.1}$ is associated with high though variable surplus production over the series, whilst F's greater than 0.4 are associated with reduced productivity in the most recent years. This analysis demonstrates the benefits of harvesting at F around $\mathrm{F}_{0.1}$. Exploitation in the range of recent $\mathrm{F}(\sim 0.7-1.2)$ is detrimental to stock productivity.

## D. Short-Term Projection

Short term forecasts were routinely performed until 2004. There was no final assessment from 2005-2008 and therefore no short term forecast was conducted. A forecast was again carried out in 2009. The method used was the "Multi fleet Deterministic Projection" software (Smith, 2000). A short-term projection is carried out under the following assumptions. Recruitment was set at geometric mean, either for the entire time series, minus the most recent two years or as in 2008, from 1995-2006. This value was around 360 million fish, and was considered a good proxy for recruitment strength in recent years. This is because the recent recruitments have fluctuated about this value. Mean weights in the catch and in the stock were calculated as means over the last three years. Selection is taken from the most recent assessment. Population number of 2 ringers in the intermediate season was calculated by the degradation of geometric mean recruitment (1995-2006) using the equation below, following the same procedure as in previous years.
$N_{t+1}=N_{t} * e^{-F_{t}+M_{t}}$

## E. Medium-Term Projections

Yield per recruit analyses have been conducted for this stock since the mid 1960s, though not necessarily every year. Recent analyses have used the "Multi Fleet Yield Per Recruit" software. A comparison of the results is shown in the table below. Based on the most recent yield per recruit $\mathrm{F}_{0.1}$ is estimated to be 0.17 (Figure 19).

Table 7 presents estimates of $\mathrm{F}_{0.1}$ from the literature and from yield per recruit analyses conducted over time. Fo. 1 estimates from the YPR analysis have been in the range 0.16-0.19. Fmax has been undefined in recent studies but earlier work suggested values of around 0.45 , based on the good recruitment regime of the 1960s.

## F. Long-Term Projections

In 2007, a number of possible management scenarios were tested using the stochastic simulation tool FPRESS (Codling and Kelly, 2005). This tool is used to test the robustness of harvest control rules.

## G. Biological Reference Points

$B_{p a}$ is set at 44000 t and $\mathrm{Blim}_{\text {lim }} 26000 \mathrm{t}$. F reference points are not defined for this stock.
$B_{p a}$ is based on a low probability of low recruitment and $B_{\lim }$ set at $B_{\text {loss. }}$ (ICES, 2001).
Reference points are defined for this stock, $\mathrm{B}_{\mathrm{pa}}$ is currently at 44000 t (low probability of low recruitment) and $B_{\lim }$ at $26000 t$ (Bloss) for this stock. $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{F}_{\text {lim }}$ are not defined. $\mathrm{F}_{\text {msy }}$ has not been estimated. However $\mathrm{F}_{0.1}$ can be assumed to be a proxy for $\mathrm{F}_{\text {msy }}$ and was estimated in 2009 to be $=0.17$.

The reference points for this stock have not been revised in recent years. There is some evidence that Blim should be revised upwards, to the point of recruitment impairment estimated by Clarke and Egan (2008). These authors showed a changepoint in a segmented regression at 47000 t .

## H.1. Biology of the species in the distribution area

Herring shoals migrate to inshore water to spawn. Their spawning grounds are located in shallow waters close to the coast and are well known and well defined. This stock can be divided into autumn and winter spawning components. Spawning begins in October and can continue until February. A number of spawning grounds are located along the South coast, extending from the Saltee Islands to the Old Head of Kinsale. These grounds include Baginbun Bay, Dunmore East Co Waterford, around Capel and Ballycotton Islands and around the entrance to Cork Harbour (Molloy, 2006). The areas surrounding the Daunt Rock and old Head of Kinsale have also been recognised as spawning grounds (Breslin, 1998). These spawning grounds are shown in Figure 2 and. 5.

Herring are benthic spawners and deposit their eggs on the sea bed usually on gravel or course sediments. The yolk sac larvae hatch and adopt a pelagic mode of life.

When referring to spawning locations the following terminology is used (Molloy, 2006)

- A spawning bed is the area over which the eggs are deposited
- A spawning ground consists of one or more spawning beds located in a small area.
- A spawning area is comprised of a number of spawning grounds in a larger area

Spawning grounds are typically located in high energy environments such as the mouth of large rivers and areas where the tidal currents are strong. Herring shoals return to the same spawning grounds each year (Molloy, 2006).

Herring produce benthic eggs that are adhered to the bottom substrate where they remain until hatching. Fertilized eggs hatch into larvae in 7-10 days depending on the water temperature ${ }^{1}$. The size of the egg determines the size of the larvae. Larger eggs have a greater chance of survival but this must be balanced against environmental conditions and the inverse relationship between fecundity and egg size (Blaxter and Hunter, 1982).

A study on fecundity of Celtic Sea herring, conducted in the 1920s found that the eggs produced by spring spawners were $25 \%$ bigger than those autumn spawners but were less numerous (Farran, 1938). Later studies of Celtic Sea herring fecundity by Molloy (1979), found that there were two spawning populations with the autumn one being most important.

The relationship between fecundity and length has been calculated for both spawning components of Celtic Sea herring. The regression equations are as shown in Hay et al 2001, are as follows:

$$
\begin{aligned}
& \text { Autumn spawning component: Fecundity }=5.1173 L-56.69(n=53) \\
& \text { Winter spawning component: Fecundity }=3.485 L-35.90(n=37)
\end{aligned}
$$

The larval phase is an important period in the herring life cycle. Larvae use their oil globule for food and to provide buoyancy. Currents transport the newly hatched larvae to areas in the Celtic Sea or to the Irish Sea (Molloy, 2006). The conditions experienced during the larval phase as well as during juvenile phase are likely to have some influence on the maturation of Celtic Sea herring. Fast growing juveniles can recruit to the population a year earlier than slow growing juveniles. Faster growth may also lead to increased fecundity (Brophy and Danilowich, 2003). Fluctuating environmental conditions play an important role in the growth and survival of herring in this area.

The juveniles tend to remain close inshore, in shallow waters for the first two years of their lives, in nursery areas. There are many of these nursery areas around the coast. The minimum landing size for herring is 20 cm and therefore these juvenile herring are not caught by the fishery in the early stages of their life cycle (Molloy, 2006).

Celtic Sea herring have undergone changes in growth patterns and a declining trend in mean weights and lengths can be seen over time. It is important to detect these changes from a management perspective because changes can have an impact on the estimation of stock size. Growth has an impact on factors such as maturity and recruitment (Molloy, 2006). Trends in mean weights and lengths are currently being examined over the time series and possible links to environmental factors investigated (Lynch in prep).

[^8]The locations of spawning and non spawning fish in the Celtic Sea as shown in Figure 5. This is based on the knowledge of fishermen and shows spawning herring are found close inshore and non spawning fish found in areas further off shore.

## H.2. Management and ICES Advice

The assessment year is from $1^{\text {st }}$ April to $31^{\text {st }}$ March. However for management purposes, the TAC year is from $1^{\text {st }}$ January to $31^{\text {st }}$ December.

The first time that management measures were applied to this fishery was during the late 1960s. This was in response to the increasing catches particularly off Dunmore East. The industry became concerned and certain restrictions were put in place in order to prevent a glut of herring in the market and a reduction in prices. Boat quotas were introduced restricting the nightly catches and the number of boats fishing. Fishing times were specified with no weekend fishing and herring could not be landed for the production of fishmeal. A minimum landing size was also introduced (Molloy, 1995).

The TAC (total allowable catch) system was introduced in 1972, which meant that yearly quotas were allocated. This continued until 1977 until the fishery was closed. During the closure a precautionary TAC was set for Division VIIj. This division was not assessed analytically (ICES, 1994). After the closure of this fishery a new management structure was implemented with catches controlled on a seasonal basis and individual boat quotas were put in place (Molloy 1995).

This fishery is still managed by a TAC system with quotas allocated to boats on a weekly basis. Participation in the fishery is restricted to licensed vessels. A series of closed areas have been implemented to protect the spawning grounds, when herring are particularly vulnerable. These spawning box closures were implemented under EU legislation.

The committee set up to manage the stock has the following objectives.

- To build the stock to a level whereby it can sustain annual catches of around $20,000 \mathrm{t}$.
- In the event of the stock falling below the level at which these catches can be sustained the Committee will take appropriate rebuilding measures.
- To introduce measures to prevent landings of small and juvenile herring, including closed areas and/or appropriate time closures.
- To ensure that all landings of herring should contain at least $50 \%$ of individual fish above 23 cm .
- To maintain, and if necessary expand the spawning box closures in time and area.
- To ensure that adequate scientific resources are available to assess the state of the stock.
- To participate in the collection of data and to play an active part in the stock assessment procedure.

The Irish Celtic Sea Herring Management Advisory Committee has developed a rebuilding plan for this stock. This Committee proposes that this plan be put forward for Council Regulation for 2009 and subsequent years. The plan incorporates scientific advice with the main elements of the EU policy statement on fishing opportunities for 2009, local stakeholder initiatives and Irish legislation.

## Proposed Rebuilding plan

1) For 2009, the TAC shall be reduced by $25 \%$ relative to the current year (2008).

2 ) In 2010 and subsequent years, the TAC shall be set equal to a fishing mortality of Fo.1.
3 ) If, in the opinion of ICES and STECF, the catch should be reduced to the lowest possible level, the TAC for the following year will be reduced by 25\%.

4 ) Division VIIaS will be closed to herring fishing for 2009, 2010 and 2011.
5 ) A small-scale sentinel fishery will be permitted in the closed area, Division VIIaS. This fishery shall be confined to vessels, of no more than 65 feet length. A maximum catch limitation of $8 \%$ of the Irish quota shall be exclusively allocated to this sentinel fishery.
6 ) Every three years from the date of entry into force of this Regulation, the Commission shall request ICES and STECF to evaluate the progress of this rebuilding plan.

7 ) When the SSB is deemed to have recovered to a size equal to or greater than $B_{p a}$ in three consecutive years, the rebuilding plan will be superseded by a long-term management plan.

Table 8 shows the history of the ICES advice, implemented TACs and ICES' estimates of removals from the stock. It can be seen that the implemented TAC has been set higher than the advice in about $50 \%$ of years since the re-opening of the fishery in 1983. The tendency for the TAC to be set higher than the advice has also increased in recent years. It can also be seen that ICES catch estimates have been lower than the agreed TAC in most years.

## H.4. Terminology

The WG uses "rings" rather than "age" or "winter rings" throughout the report to denominate the age of herring, with the intention to avoid confusion. It should be observed that, for autumn spawning stocks, there is a difference of one year between "age" and "rings". HAWG in 1992 (ICES 1992/Assess:11) stated that
"The convention of defining herring age rings instead of years was introduced in various ICES working groups around 1970. The main argument to do so was the uncertainty about the racial identity of the herring in some areas. A herring with one winter ring is classified as
 the age of the herring in rings instead of in years allowed scientists to postpone the decision on year of birth until a later date when they might have obtained more information on the racial identity of the herring.

The use of winter rings in ICES working groups has introduced a certain amount of confusion and errors. In specifying the age of the herring, people always have to state explicitly whether they are talking about rings or years, and whether the herring are autumn- or spring spawners. These details tend to get lost in working group reports, which can make these reports confusion for outsiders, and even for herring experts themselves. As the age of all other fish species (and of herring in other parts of the world) is expressed in years, one could question the justification of treating West-European herring in a special way. Especially with the present trend towards multispecies assessment and integration of ICES working groups, there might be a case for a uniform system of age definition throughout all ICES working groups.

However, the change from rings to years would create a number of practical problems. Data files in national laboratories and at ICES would have to be adapted, which would involve extra costs and manpower. People that had not been aware of the change might be confused when comparing new data with data from old working group reports. Finally, in some areas (notably Division IIIa), the distinction between spring- and autumn spawners is still hard to make, and scientists preferred to continue using rings instead of years.

The Working Group discussed at length the various consequences of a change from rings to years. The majority of the Group felt that the advantages of such a change did not outweigh the disadvantages, and it was decided to stick to the present system for the time being."

The text table below gives an example for the correlation between age, rings and year class for the different spawning types in late 2002:

| YEAR CLASS (AUTUMN SPAWNERS) | 2001/2002 | $\mathbf{2 0 0 0 / 2 0 0 1}$ | 1999/2000 | 1998/1999 |
| :--- | :--- | :--- | :--- | :--- |
| Rings | 0 | 1 | 2 | 3 |
| Age (autumn spawners) | 1 | 2 | 3 | 4 |
| Year class (spring spawners) | 2002 | 2001 | 2000 | 1999 |
| Rings | 0 | 1 | 2 | 3 |
| Age (spring spawners) | 0 | 1 | 2 | 3 |

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Figure 1. Herring in the Celtic Sea and VIIj. Schematic presentation of the life cycle of Celtic Sea and VIIj Herring (ICES, 2005c, SGRESP).


Figure 2. Herring in the Celtic Sea and VIIj. Schematic presentation of prevailing oceanographic conditions in the Celtic Sea and VIIj (ICES, 2005c, SGRESP).


Figure 3. Herring in the Celtic Sea and VIIj. Areas mentioned in the text and spawning boxes A, $B$ and C, south of Ireland. One of these boxes is closed each season, under EU legislation. 1 Courtmacsherry, 2 Cork Harbour, 3 Daunt Rock, 4 Kinsale Gas Field (Rigs), 5 Labadie Bank, 6 Kinsale, 8 Waterford Harbour, 9, Baginbun Bay, 10, Tramore Bay/ Dunmore East, 11, Ballycotton Bay, 12, Valentia Island, 13 Kerry Head to Loop Head, 14, The Smalls. The spawning boxes A-C correspond to ICES Divisions VIIj, VIIg and VIIaS respectively.


Figure 4. Herring in the Celtic Sea and VIIj. Spawning ground of herring along the south coast of Ireland, inferred from information on the Irish herring fishery (Breslin, 1998).


Figure 5. Herring in the Celtic Sea and VIIj. Location of spawning (closed symbol) and non spawning (open symbol) herring in the Celtic Sea and SW of Ireland, based on expert fishemens' knowledge.


Figure .6. Herring in the Celtic Sea and VIIj. ICES estimates of herring catches (tonnes) per season 1958/1959 to 2008/2009.


Figure 7. Herring in the Celtic Sea and VIIj. Catch numbers at age standardised by yearly mean.


Figure 8. Herring in the Celtic Sea and VIIj. Trends over time in mean weights in the catch.


Figure 9. Herring in the Celtic Sea and VIIj. Trends over time in mean weights in the stock at spawning time.



Figure 10. Herring in the Celtic Sea and VIIj. Log catch ratios (above) and log catch ratios smoothed with a 4 year moving average for each age group for the time series 1958-2006. Evidence of a change in selection pattern visible in upper panel in 2003.


Figure 11. Herring in the Celtic Sea and VIIj. Cohort catch curves for the time series of catch at age data. Age in winter rings on the horizontal axis and log transformed catch numbers at age on the vertical axis.


Figure 12: Herring in the Celtic Sea and VIIj. Total mortality (Z) estimated from cohort catch curves (2-7 ringer) for cohorts 1958 to 1997.


Figure 13. Herring in the Celtic Sea and VIIj. Cohort catch curves (2-5 ringer), averaged over several year classes, from catch at age data.


Figure 14. Herring in the Celtic Sea and VIIj. Cohort catch curves (2-5 ring) based on acoustic survey abundance. Upper panel shows means for two periods, and below for three time periods, over the same series of surveys


Figure 15. Herring in the Celtic Sea and VIIj. SSB, F and recruitment (1-ringer) from proposed final run. Note SSB in the terminal year is adjusted according to the protocol for this stock.


Figure 16. Herring in the Celtic Sea and VIIj. Stock recruit relationship from ICA base case runs. Data classified according to quality of input data, see Table 4.


Figure 17. Herring in the Celtic Sea and VIIj. Recruits per spawner, in '000s/tonnes


Figure 18. Herring in the Celtic Sea and VIIj. Total and surplus production in the time series over a range of fishing mortalities.



Figure 19. Herring in the Celtic Sea and VIIj. Yield per recruit carried out in 2009

| Reference point | F multiplier | Absolute <br> F |
| :---: | :---: | :---: |
| Fbar(2-5) | 1 | 0.1253 |
| FMax | $>=1000000$ |  |
| F0.1 | 1.3466 | 0.1688 |
| F35\%SPR | 1.5231 | 0.1909 |

Table 1. Herring in the Celtic Sea \& Division VIIj. Acoustic surveys of Celtic Sea and VIIj herring, by season. Number of surveys per season and type indicated along with biomass and SSB estimates. Shaded sections show surveys not used in tuning, in most recent assessment.

| Season | No. | Type | Survey Timing | SSB |
| :---: | :---: | :---: | :---: | :---: |
| 1990/1991 | 2 | Autumn and winter spawners | Oct and Jan/Feb | - |
| 1991/1992 | 2 | Autumn and winter spawners | Nov/Dec and Jan | - |
| 1992/1993 | 2 | Autumn and winter spawners | Nov and Jan | - |
| 1993/1994 | 2 | Autumn and winter spawners | Nov and Jan | - |
| 1994/1995 | 2 | Autumn and winter spawners | Nov and Jan | - |
| 1995/1996 | 2 | Autumn and winter spawners | Nov and Jan | 36 |
| 1996/1997 | 1 | Autumn and winter spawners | Oct/Nov and Jan | 151 |
| 1997/1998 | - | No survey |  | - |
| 1998/1999 | 1 | Autumn spawners | Nov and Jan | 100 |
| 1999/2000 | 1 | Feeding phase | July | - |
| 1999/2000 | 1 | Winter-spawners | Nov and Jan | - |
| 2000/2001 | 2 | Autumn and winter spawners | Oct and Jan | 20 |
| 2001/2002 | 2 | Pre-spawning | Sept and Oct | 95 |
| 2002/2003 | 1 | Pre-spawning | Sept/Oct | 41 |
| 2003/2004 | 1 | Pre-spawning | Oct/Nov | 20 |
| 2004/2005 | 1 | Pre-spawning | Nov/Dec | - |
| 2005/2006 | 1 | Pre-spawning | Oct | 33 |
| 2006/2007 | 1 | Pre-spawning | Oct | 36 |
| 2007/2008 | 1 | Pre-spawning | Oct | 46 |
| 2008/2009 | 1 | Pre-spawning | Oct | 90 |

Table 2. Herring in the Celtic Sea \& Division VIIj. Original acoustic survey abundance at age as used by ICES until HAWG 2006.


Table 3. Herring in the Celtic Sea \& Division VIIj. Revised acoustic series as used by HAWG since 2006. Shaded colums show surveys excluded from tuning in 2009, where timing and design of earlier surveys were not considered comparable with the sufficiently long series of subsequent surveys.

| 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 202 | 3 | - | 0 | - | 25 | 40 | 0 | 24 | - | 2 | - | 1 | 99 |
| 25 | 164 | - | 30 | - | 102 | 28 | 42 | 13 | - | 65 | 21 | 106 | 64 |
| 157 | 795 | - | 186 | - | 112 | 187 | 185 | 62 | - | 137 | 211 | 70 | 295 |
| 38 | 262 | - | 133 | - | 13 | 213 | 151 | 60 | - | 28 | 48 | 220 | 111 |
| 34 | 53 | - | 165 | - | 2 | 42 | 30 | 17 | - | 54 | 14 | 31 | 162 |
| 5 | 43 | - | 87 | - | 1 | 47 | 7 | 5 | - | 22 | 11 | 9 | 27 |
| 3 | 1 | - | 25 | - | 0 | 33 | 7 | 1 | - | 5 | 1 | 13 | 6 |
| 1 | 15 | - | 24 | - | 0 | 24 | 3 | 0 | - | 1 | - | 4 | 5 |
| 2 | 0 | - | 4 | - | 0 | 15 | 0 | 0 | - | 0 | - | 1 |  |
| 2 | 2 | - | 2 | - | 0 | 52 | 0 | 0 | - | 0 | - | 0 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  | - |
| 469 | 1338 | - | 656 |  | 256 | 681 | 423 | 183 | - | 312 | 305 | 454 | 769 |
| 36 | 151 |  | 100 |  | 20 | 95 | 41 | 20 | - | 33 | 36 | 46 | 90 |
| 53 | 26 |  | 36 |  | 100 |  | 49 | 34 | - | 48 | 35 | 25 | 20 |
| AR | AR |  | AR |  | AR | AR | AR | AR |  | $R$ | $R$ | $R$ | $R$ |

Table 4. Herring in the Celtic Sea \& Division VIIj. Rudimentary history of the Irish fishery since 1958.

| Time period | 1958-1977 | 1977-1983 | 1983-1997 | 1998-2004 | 2004-2007 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Type of fishery | Cured fish | Closure | Herring roe | Fillet/whole fish | Fillet/whole fish |
| Quality of catch data | High | Medium | Low | Medium/low | High |
| Source of catch data | Auction data | Auction data | Skipper logbook estimate | Skipper logbook estimate | Weighbridge landings |
| Discard Levels | Low | Low | High | Medium | Medium |
| Incentive to discard | None | None | Maturity stage | Size grad | market vs. quota |
| Alloowance for water* | na | na | na | 20\%* | 2\%* |

* RSW only. These vessels are more dominant in recent years.

Table 5. Celtic Sea and VIIj herring. Biological history of the stock.

|  | $1958-1972$ | $1973-1977$ | $1978-1980$ | $1981-1983$ | $1984-1995$ | $1996-2008$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |
| MW 2-ring (kg) <br> median | 0.146 | 0.181 | 0.179 | 0.158 | 0.135 | 0.115 |
| ML 2-ring (cm) median 26.4 | 27.5 | 27.1 | 26.3 | 25.2 | 24.4 |  |
| Z (cohort catch curve) | $0.22-0.93$ | $0.42-1.12$ | $0.74-0.93$ | $0.62-0.74$ | $0.49-0.89$ | $0.48-1.01$ |
| GM recruitment 106 | 448 | 167 | 168 | 587 | 514 | 340 |
| Recruitment anomaly | positive | negative | negative | positive | positive | both |
| SSB (000 t) | $53-126$ | 27 to 52 | $25-26$ | $30-63$ | $49-68$ | $24-70$ |
| F (2-5 r) | $0.23-0.71$ | $0.55-0.80$ | $0.50-0.68$ | $0.68-0.87$ | $0.40-0.98$ | $0.12-0.88$ |

Table 6. Celtic Sea and VIIj herring. Total mortality Z estimated from cohort catch curves.

| Cohort | Z (2-7 ring) | Cohort | Z (2-7 ring) |
| :--- | :---: | :--- | :--- |
|  |  |  |  |
| 1956 | 0.39 | 1977 | 1.09 |
| 1957 | 0.37 | 1978 | 0.84 |
| 1958 | 0.31 | 1979 | 0.93 |
| 1959 | 0.42 | 1980 | 0.75 |
| 1960 | 0.22 | 1981 | 0.75 |
| 1961 | 0.47 | 1982 | 0.65 |
| 1962 | 0.30 | 1983 | 0.63 |
| 1963 | 0.50 | 1984 | 0.50 |
| 1964 | 0.62 | 1985 | 0.66 |
| 1965 | 0.71 | 1986 | 0.62 |
| 1966 | 0.66 | 1987 | 0.76 |
| 1967 | 0.51 | 1988 | 0.58 |
| 1968 | 0.93 | 1989 | 0.73 |
| 1969 | 0.82 | 1990 | 0.57 |
| 1970 | 0.76 | 1991 | 0.65 |
| 1971 | 0.55 | 1992 | 0.77 |
| 1972 | 0.51 | 1993 | 0.90 |
| 1973 | 0.43 | 1994 | 0.73 |
| 1974 | 0.68 | 1995 | 0.80 |
| 1975 | 0.86 | 1996 | 1.02 |
| 1976 | 1.12 | 1997 | 0.88 |
|  |  |  |  |

Table 7. Celtic Sea and VIIj herring. Estimates of estimates of $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\text {max }}$ from the literature and HAWG work.

|  | F0.1 | $\mathrm{F}_{\text {max }}$ | MSY | Comments | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | - | >0.5 | $\begin{aligned} & 12- \\ & 15 \\ & 000 t \end{aligned}$ | Years for calculation had lower recruitment | Burd and Bracken, 1965 |
|  |  |  | 22 |  |  |
| 1969 | - | $\sim 0.45$ | 000 t | Years for calculation had higher recruitment | Molloy, 1969 |
|  |  |  | 14 |  |  |
| 1974 | - | >0.5 | 000 * | Fmsy calculated for periods of high and low recruitment | Corten, 1974 |
| 1983 | 0.16 |  |  | Yield/Biomass ratio | HAWG, 1983 |
| 1990 | 0.16 |  |  |  | HAWG, 1990 |
| 1994 | 0.16 |  |  |  | HAWG, 1994 |
| 1995 | 0.16 |  |  |  | HAWG, 1995 |
| 1996 | 0.16 |  |  |  | HAWG, 1996 |
| 1997 | 0.1 |  |  |  | HAWG, 1997 |
| 1999 | <0.2 |  |  |  | HAWG, 1999 |
| 2000 | <0.2 |  |  |  | HAWG, 2000 |
| 2002 | 0.17 |  |  | MFYPR software | HAWG, 2002 |
| 2003 | 0.17 |  |  | MFYPR software | HAWG, 2003 |
| 2004 | 0.17 |  |  | MFYPR software | HAWG, 2004 |
| 2007 | 0.19 |  |  | MFYPR software | HAWG, 2007 |
| 2009 | 0.17 |  |  | MFYPR software | HAWG 2009 |

*endorses Molloy (1969) provided that recruitment is at level 1966-1969

Table 8 Celtic Sea and VIIj herring. Advice history.

| ICES <br> Advice | Predicted catch corresp. to advice | Agreed TAC | Official <br> Landings | Discards | Estimated Catch ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| NEAFC TAC |  | 32 | 20 | - | 19.74 |
| Reduce F, TAC ? 25,000 |  | 25 | 16 | - | 15.13 |
| TAC between 10,000 and 12,000 |  | 10.8 | 10 | - | 8.2 |
| No Fishing | 0 | 0 | 8 | - | 3.0 |
| No Fishing | 0 | 0 | 8 | - | 7.1 |
| TAC set for VIIj only, No fishing in Celtic Sea | 0 | 6 | 10 | - | 12.1 |
| TAC set for VIIj only, No fishing in Celtic Sea |  | 6 | 9 | - | 9.2 |
| TAC set for VIIj only, No fishing in Celtic Sea |  | 6 | 17 | - | 16.8 |
| TAC |  | 8* | 10 | - | 9.5 |
| TAC |  | 8* | 22 | 4.0 | 22.18 |
| TAC | 13 | 13 | 20 | 3.6 | 19.7 |
| TAC | 13 | 13 | 16 | 3.1 | 16.23 |
| No specific TAC, preferred overall catch 17,000t |  | 17 | 13 | 3.9 | 23.3 |
| Precautionary TAC | 18 | 18 | 18 | 4.2 | 27.3 |
| TAC | 13 | 18 | 17 | 2.4 | 19.2 |
| TAC | 20 | 20 | 18 | 3.5 | 22.7 |
| TAC | 15 | 17.5 | 17 | 2.5 | 20.2 |
| TAC (TAC excluding discards) | 15 (12.5) | 21 | 21 | 1.9 | 23.6 |
| TAC | 27 | 21 | 19 | 2.1 | 23 |
| Precautionary TAC (including discards) | 20-24 | 21 | 20 | 1.9 | 21.1 |
| Precautionary TAC (including discards) | 20-24 | 21 | 19 | 1.7 | 19.1 |
| No specific advice | - | 21 | 18 | 0.7 | 19 |
| TAC | 9.8 | 16.5-21 | 21 | 3 | 21.8 |
| If required, precautionary TAC | <25 | 22 | 20.7 | 0.7 | 18.8 |
| Catches below 25 | <25 | 22 | 20.5 | 0 | 20.3 |
| $\mathrm{F}=0.4$ | 19 | 21 | 19.4 | 0 | 18.1 |
| $\mathrm{F}<0.3$ | 20 | 21 | 18.8 | 0 | 18.3 |
| $\mathrm{F}<0.34$ | 17.9 | 20 | 19 | 0 | 17.7 |
| F<0.35 | 11 | 11 | 11.5 | 0 | 10.5 |
| Substantially less than recent catches | - | 13 | 12 | 0 | 10.8 |
| 60\% of average catch 1997-2000 | 11 | 13 | 12 | - | 11 |
| 60\% of average catch 1997-2000 | 11 | 13 | 10 | - | 8 |
| Further reduction 60\% avg catch 2002-2004 | 6.7 | 11 | 9 | - | 8.5 |
| No fishing without rebuilding plan | -- | 9.3 | 9.6 | - | 8.2 |
| No targeted fishing without rebuilding plan | -- | 7.9 | 7.8 |  | 6.7 |
| No targeted fishing without rebuilding plan | -- | 5.9 |  |  |  |

* TAC from $1^{\text {st }}$ Oct $-31^{\text {st }}$ Mar

1) Calendar year

## Annex 6 - Stock Annex Herring in VlaN

Quality Handbook ANNEX: Hawg-her47d3
Stock specific documentation of standard assessment procedures used by ICES.

Stock: Herring in VIa (North)<br>Working Group: Herring Assessment WG for the Area south of $62^{\circ} \mathrm{N}$<br>Date:<br>25 March 2008<br>Authors:<br>E.M.C. Hatfield, E.J. Simmonds and A. Edridge

## A. General

## A.1. Stock definition

The stock is distributed over ICES Division VIa (N). Some of the larger adults typically found close to the shelf break may be caught in division Vb .

## A.2. Fishery

The dominant fleet fishing in VIa (N) since 1957 has been the Scottish fleet. In the early years the Scottish fishery was prosecuted using a mixture of vessel size and gear, including gill nets, ring-nets and trawls. The boats were small, and targeted the coastal stock, primarily fishing in the winter. Until 1970 the only other nations fishing in this area on a regular basis were the former German Federal Republic, and to a much lesser extend the Netherlands. These fleets operated in deeper water near the shelf edge.

In 1970 a large increase in exploitation occurred with the entry of fleets from Norway and the Faroes, and an increased Netherlands catch. In addition, considerably smaller catches were taken by France and Iceland.

Throughout this period juvenile herring catches from the Moray Firth, in the northeast of Scotland, were included in the VIa catch figures, as tagging programs showed there to be some links between herring spawning to the west of Scotland and the Moray Firth juveniles.

Prior to 1982 herring stocks in ICES Area VIa were assessed as one stock, along with the herring by-catch from the sprat fishery in the Moray Firth. In the 1982 herring assessment working group report, and in subsequent years, Area VIa was split into a northern and a southern area at $56^{\circ} \mathrm{N}$ (ICES, 1982).

In 1979 and 1981 the fishery was closed. After re-opening the nature of the fishery changed to an extent, with fewer Scottish boats targeting the coastal stock than before the closure. The Scottish domestic pair trawl fleet and the Northern Irish fleet operated in shallower, coastal areas, principally fishing in the Minches and around the Island of Barra in the south; younger herring are found in these areas. Since 1986 Irish trawlers have operated in the south of the area, from the VIa (S) line up to the southwestern Hebrides. The Scottish and Norwegian purse seine fleets targeted herring mostly in the northern North Sea, but also operated in the northern part of VIa (N). An international freezer-trawler fishery operated in deeper water near the shelf edge where older fish are distributed. These vessels are mostly registered in the Nether-
lands, Germany, France and England. In recent years the catch of these fleets has become more similar and has been dominated by younger adults resulting from increased recruitment into the stock.

In recent years the Scottish fleet has changed to a predominantly purse-seine fleet to a trawl fleet. Norwegian vessels fish less in the area than in the past. Scottish catches still comprise around half of the total, the rest is dominated by the offshore, international fishery.
A recent EU-funded programme WESTHER has elucidated stock structures of herring throughout the western seaboard of the British Isles using a combination of morphometric measurements, otolith structure, genetics and parasite loads. The results provide information on mixing of stocks within and beyond VIa (N).

## A.3. Ecosystem aspects

Herring are an important prey species in the ecosystem and also one of the dominant planktivorous fish.
Herring fisheries tend to be clean with little bycatch of other fish. Scottish discard observer programs since 1999 indicate that discarding of herring in these directed fisheries are at a low level. These discard observer programs have recorded occasional catches of seals and zero catches of cetaceans.

## B. Data

## B.1. Commercial catch

Commercial catch is obtained from national laboratories of nations exploiting herring in VIa (N). Since 1999 (catch data 1998), these labs have used a spreadsheet to provide all necessary landing and sampling data, which was developed originally for the Mackerel Working Group (WGMHSA) and further adapted to the special needs of the Herring Assessment Working Group. The current version used for reporting the 2002 catch data was v1.6.4. The majority of commercial catch data of multinational fleets was provided on these spreadsheets and further processed with the SALLOCLapplication (Patterson, 1998a). This program gives the needed standard outputs on sampling status and biological parameters. It also clearly documents any decisions made by the species co-ordinators for filling in missing sampling data and raising the catch information of one nation/quarter/area with information from another data set.

Transparency of data handling by the Working Group. The current practice of data handling by the Working Group is that the data received by the co-ordinators is available in a folder called "archive". These high-resolution data are not reproduced in the report. The archived data contains the disaggregated dataset (disfad), the allocations of samples to unsampled catches (alloc), the aggregated dataset (sam.out) and (in some cases) a document describing any problems with the data in that year.

Current methods of compiling fisheries assessment data. The species co-ordinator is responsible for compiling the national data to produce the input data for the assessments. In addition to checking the major task involved is to allocate samples of catch numbers, mean length and mean weight-at-age to unsampled catches. There are at present no defined criteria on how this should be done, but the following general process is implemented by the species co-ordinators. Searches are made for appropriate samples by gear (fleet) area quarter, if an exact match is not available the search will move to a neighbouring area if the fishery extends to this area in the same quarter. More than one sample may be allocated to an unsampled catch, in this case a
straight mean or weighted mean of the observations may be used. If there are no samples available the search will move to the closest non-adjacent area by gear (fleet) and quarter, but not in all cases.

Until 2003 the VIa(N) catch data extended back to the early 1970s; since 1986 the series has run from 1976 to present. In 2004 the data set was extended back to 1957. Details are given below.

## Historic Catches from 1957 to 1975

The working group has obtained preliminary estimates of catch and catch-at-age for the period 1957 to 1975 . These have been estimated from records of catch presented in HAWG reports from 1973, 1974, 1981 and 1982. Intervening reports were also consulted to check for changes or updates during the period. Catch-at-age data were available from 1970 to 1975 from the 1982 Working Group report, and catches-at-age for the period 1957 to 1972 were estimated from paper records of catch-at-age by national fleets for 1957 to 1972, held at FRS Marine Laboratory Aberdeen. The fishing practices of national fleets were established for the period 1970 to 1980 from catches in VIa and VIa (N) recorded in the 1981 and 1982 Working Group reports respectively. This procedure suggested that, on average, more than $90 \%$ of catch by national fleet could be fully assigned to either VIa (N) or VIa (S). The remaining catch was assigned assuming historic proportions. During this period catches were split into autumn and spring spawning components; anecdotal information on trials to verify this separation suggests it was not a robust procedure. Currently about $5 \%$ of herring in VIa $(\mathrm{N})$ is found to be spent at the time of the acoustic surveys in July, and thought to be spring spawning herring. However, at present the Working Group assesses VIa $(\mathrm{N})$ herring as one stock, regardless of spawning stock affiliation. In the earlier period higher proportions were allocated as spring spawners. Currently the designated 'spring spawning' component is not included in the catch at age matrix, but the catch tones express the full amount giving rise to SoP differences in the early years. Similarly, a small Moray Firth juvenile fishery was also included in VIa (N) catch in earlier years because it was thought that these juveniles were part of the VIa ( N ) stock. Separating this component in the historic data was difficult, and as the fishery ceased in the very early 70s this has no implications for current allocation of these fish. The Moray Firth is, geographically, part of IVa (ICES stat. rectangles 44E6, 44E7, 45E6) and is now managed as part of that area. Currently there are no juvenile herring catches from the Moray Firth. Full details of the analysis carried out is provided as an appendix (Appendix 11) to the 2004 Working Group report. Further investigations are required before determining the correct actions concerning the 'spring spawners' in early period. The consequence of this is to slightly reduce the apparent stock size in the early years, when is already at an all time high. It has no implications for fitting of any survey data, or influence on the Blim reference point, however, it might further increase the high $R$ seen at high $S S B$ in a $S / R$ relationship.

## Allocation of catch and misreporting

This fishery has had a strong tradition of misreporting before 2000, though this has reduced in recent years. It is believed that the shortfall between the TAC and the catch was used to misreport catches from other areas (from IVa to the east and from VIa (S) to the south). In the past, fishery-independent information confirmed that large catches were being reported from areas with low abundances of fish, and informal information from the fishery and from other sources confirmed that most catches of fish recorded between $4{ }^{\circ} \mathrm{W}$ and $5^{\circ} \mathrm{W}$ were most probably misreported North Sea catches. The problem was detailed in the Working Group report in 2002
(ICES 2002/ACFM:12). Improved information from the fishery in 1998-2002 allowed for re-allocation of many catches due to area misreporting (principally from VIa (N) to IVa (W)). This information was obtained from only some of the fleets
As a result of perceived problems of area misreporting of catch from IVa into VIa (N), Scotland introduced a fishery regulation in 1997 with the aim to improve reporting accuracy. Under this regulation, Scottish vessels fishing for herring were required to hold a license either to fish in the North Sea or in the west of Scotland area (VIa (N)). Only one licensed option could be held at any one time. However in 2004, the requirement to carry only a single licence was rescinded. Area misreporting of catch taken in area IVa into area VIa (N) then increased in 2004 and continued in 2005. It is possible, therefore, that the relaxation of this single area licence contributed to a resurgence in area misreporting. In 2007, as in 2006, there was no misreporting from IVa into VIa (N). New sources of information on catch misreporting from the UK became available in 2006 (see the 2007 HAWG report). This information was associated with a stricter enforcement regime that may be responsible for the lack of that area misreporting since 2006.

The Butt of Lewis box, (a seasonal closure to pelagic fishing of the spawning ground in the north west of the continental shelf in area VIa(North) since the late 1970s was opened to fishing in 2008 following a STECF review in 2007. It has not been possible to show either beneficial or deleterious effects from this closure.

Catches are included in the assessment. Biases and sampling designs are not documented. Discards are not included, though data from some fleets suggest these are very minor. Slippage and high grading are not recorded.

## B.2. Biological

Catch-at-age data (catch numbers-at-age, mean weights-at-age in the catch, mean length-at-age) are derived from the raised national figures received from the national laboratories. The data are obtained either by market sampling or by onboard observers, and processed as described in Section B. 1 above. For information on recent sampling levels and nations providing samples, see Section 2.2. in the most recent HAWG report.
Proportions mature (maturity ogive) and mean weights-at-age in the stock derived from the acoustic survey (see next section) have been used since 1992 and 1993, respectively. Prior to these years, time-invariant values derived from ??? were used.

Biological sampling of the catches was extremely poor in recent history (particularly in 1999). This was particularly the case for the freezer trawler fishery that takes the larger component of the stock based around the shelf break. The lack of samples was due in part to the fact that national vessels tend to land in foreign ports, avoiding national sampling programs. The same fleet is thought to high grade. The long length of fishing trips makes observer programs difficult. Even when samples are taken, age determination is limited for most nations.
Sampling has improved over the last few years. The number of age readings per 1,000 t of catch increased from the low in 1999 of 52 to a high in 2001 of 93 . Numbers have decreased again since then to 57 per 1,000 t in 2003. From 1999 to 2003 the sampling has been dominated by Scotland (ranging between 70 and $98 \%$ of the age readings), except in 2001, when only $43 \%$ of the age determination was on Scottish landings in VIa (N).

Natural mortality (M) varies with age (expressed in number of winter rings) according to the following:

Rings M

| 1 | 1 |
| :--- | :--- |
| 2 | 0.3 |
| 3 | 0.2 |
| $4+$ | 0.1 |

Those values have been held constant from 1957 to date. Those values correspond to estimates for North Sea herring based on recommendations by the Multi-species WG (Anon. 1987a) that were applied to adjacent areas (Anon. 1987b).

## B.3. Surveys

## B.3.1 Acoustic survey

An acoustic survey has been carried out for VIa (N) herring in the years 1987, 19912003

Biomass estimated from the acoustic survey tends to be variable. Herring are found in similar area each year, namely south of the Hebrides off Barra Head, west of the Hebrides and along the shelf edge.

The stock is highly contagious in its spatial distribution, which explains some of the high variability in the time series. Effort stratification has improved with knowledge of the distribution and this may be less of a problem in more recent years. The survey uses the same target strength as for the North Sea surveys and there is no reason to suppose why this should be any different. Species identification is generally not a great problem.

## B.3.2 Larvae survey

Larvae surveys for this stock were carried out from 1973 to 1993. Larval production estimates (LPE) and a larval abundance index (LAI) were produced for the time series. These values were used in the assessment, the LPE until 2001. However, in 2002 it was decided that the LAI had no influence on the assessment and has not been used since. Documentation of this survey time-series is given in ICES CM 1990/H:40.

## B.4. Commercial CPUE

Not used for pelagic stocks

## B.5. Other relevant data

## C. Historical Stock Development

An experimental survey-data-at-age model was formulated at the 2000 HAWG. In 1999 and 1998 a Bayesian modification to ICA was used to account for the uncertainty in misreporting.

Model used: FLICA Software R / ICA (Patterson 1998b)
Model Options chosen:
Separable constraint over last 8 years (weighting $=1.0$ for each year)

Reference age $=4$
Constant selection pattern model
Selectivity on oldest age $=1.0$
First age for calculation of mean $F=3$
Last age for calculation of mean $\mathrm{F}=6$
Weighting on 1-rings $=0.1$; all other age classes $=1.0$
Weighting for all years $=1.0$
All indices treated as linear
No S/R relationship fitted
Lowest and highest feasible F $=0.02$ and 0.5
All survey weights equal i.e., 1.0 with the exception of 1 ringers in the acoustic survey weighted to 0.1.

Correlated errors assumed i.e., $=1.0$
No shrinkage applied
Input data types and characteristics:

| Type | Name | Year range | Age range | Variable from year to year Yes/No |
| :---: | :---: | :---: | :---: | :---: |
| Caton | Catch in tones | 1957 - last data year | NA | Yes |
| Canum | Catch at age in Numbers | 1957 - last data year | 1-9+ | Yes |
| Weca | Weight at age in the commercial catch | $\begin{aligned} & \text { 1957-1972 1973- } \\ & \text { 1981 1982-1984 } \\ & \text { 1985-last data year } \end{aligned}$ | $\begin{aligned} & 1-9+1-9+1-9+1- \\ & 9+ \end{aligned}$ | No <br> No No Yes |
| West | Weight at age of the spawning stock at spawning time. | $\left\lvert\, \begin{aligned} & 1957-1992 \\ & \text { 1993-last data year } \end{aligned}\right.$ | $\left\lvert\, \begin{aligned} & 1-9+ \\ & 1-9+ \end{aligned}\right.$ | No Yes |
| Mprop | Proportion of natural mortality before spawning | 1957-last data year | NA | No |
| Fprop | Proportion of fishing mortality before spawning | 1957-last data year | NA | No |
| Matprop | Proportion mature at age | $\begin{aligned} & 1957 \text { - } 1991 \\ & \text { 1992-last data year } \end{aligned}$ | $\begin{array}{\|l} 1-9+ \\ 1-9+ \end{array}$ | No <br> Yes |
| Natmor | Natural mortality | 1957 - last year | 1-9+ | No |

Tuning data:

| Type | Name | Year Range | Age Range |
| :--- | :--- | :--- | :--- |
| Tuning fleet 1 | VIa (N) Acoustic Survey | 1987, | $1-9+$ |
|  |  | 1991 - last data year | $1-9+$ |

## D. Short-Term Projection

Model used: Age structured Software used: MFDP ver 1a
Initial stock size: Taken from the last year of the assessment. 1- and 2-ring recruits taken from a geometric mean for the years 1976 to one year prior to the last year.

Maturity: Mean of the last three years of the maturity ogive used in the assessment.
F and M before spawning: Set to 0.67 for all years.
Weight at age in the stock: Mean of the last three years in the assessment.
Weight at age in the catch: Mean of the last three years in the assessment.
Exploitation pattern: Mean of the previous eight years, scaled by the Fbar (3-6) to the level of the last year (eight because this is the assessment model assumption of 8 years separable period).

Intermediate year assumptions: TAC constraint. Stock recruitment model used: None used

Procedures used for splitting projected catches: Not relevant

## E. Medium-Term Projections (done intermittently)

Model used: STPR as described in Skagen (2003)
Initial stock size: Population parameters Terminal year survivors from ICA assessment with recruits replaced as in short term projections (D above). Drawn from a multivariate lognormal distribution with mean equal to the values estimated in the stock assessment model, and with covariance as estimated in the same model fit. Geometric mean recruitment for 1 - and 2-ringers is used to replace the values in the assessment for the first projected year, covariance at age 2 retained and used for age 1 and 2.

Natural mortality: Mean of the last three years in the assessment.
Maturity: drawn randomly by year from 1990 to present.
F and M before spawning: Set to 0.67 for all years.
Weight at age in the stock: drawn randomly by year from 1990 to present.
Weight at age in the catch: drawn randomly by year from 1990 to present.
Exploitation pattern: from the eight year separable model
Intermediate year assumptions: TAC constraint
Stock recruitment model used: Variable Hockey-Stick or Beverton Holt fitted to recent data (1989 on), but other options tested for robustness max year three years prior to the assessment.

## G. Biological Reference Points

The report of SGPRP (ICES 2003/ACFM:15) proposed a Blim of 50,000 t for VIa (N)
herring. This is calculated from the values in the converged part of the VPA (19761999) and the Working Group endorsed this value in 2003 (ICES 2003/ACFM:17).

Suggested Precautionary Approach reference points:

| $\mathrm{B}_{\mathrm{LIM}}$ is $50,000 \mathrm{t}$ | $\mathrm{B}_{\mathrm{PA}}$ be set at $75,000 \mathrm{t}$ |
| :--- | :--- |
|  |  | | Technical basis: <br> $\mathrm{B}_{\text {LIM }}: \mathrm{B}_{\text {Loss }}$ Estimated SSB for sustained <br> recruitment | Bpa: $1.5 *$ Blim |
| :--- | :--- |
|  |  |

## H. Other Issues

## H. 1 Biology of the species in the distribution area

The Atlantic herring, Clupea harengus, is numerically one of the most important pelagic species in North Atlantic ecosystems with widespread distribution around the Scottish coast. Within the Northeast Atlantic they are encountered from the north of Biscay to Greenland, and east into the Barents Sea. It is thought that herring stocks comprise many reproductively isolated subpopulations through specific spawning grounds and seasons (e.g. autumn and spring spawners), but the taxonomic status of these subpopulations remains unclear.

Herring are demersal spawners and produce dense beds of benthic eggs deposited on gravelly substrates. This behaviour is considered to be an evolutionary remnant of herrings' river spawning past. Each female produces a single batch of eggs per year, releasing a ribbon of eggs that adheres to the benthos; the male sheds milt while swimming a few centimetres above the female. This particular behaviour renders herring vulnerable to anthropogenic activity such as offshore oil and gas industries and gravel extraction.
The eggs take about three weeks to hatch, dependant on the temperature. The larvae on hatching are $6-9 \mathrm{~mm}$ long and are immediately planktonic. Their yolk sac lasts for about a week during which time they will begin to feed on phytoplankton and crustacean larvae. Their planktonic development lasts around three to four months during which time they are passively subjected to the residual drift which takes them to coastal nurseries. The habitats of juveniles are primarily pelagic, and hydrographical features such as temperature and the depth of thermocline, as well as abundance of zooplankton affect their distribution. Adult fish are pelagic and found mostly in continental shelf seas to depths up to 200 m . They form large shoals with diurnal migration patterns through the water column which can be associated with the availability of prey and stage of maturity. In the winter the feeding activity and growth are very slow. Herring can reach 40 cm in length and have a maximum lifespan of 10 years although most herring range between $20-30 \mathrm{~cm}$ and are less than 7 years.

Assessing age and year class for herring can be problematic due to the extended spawning season of autumn spawners from September to January. Using the convention of January $1^{\text {st }}$ as the birthday, 0-group refer to fish born between 3 and 18 months ago but 0-group autumn spawners belong to a different class from 0-group spring spawners. Time series of a stock's age structure helps its management and it is vital that they are extended for all the 'West of Scotland' herring components in the

VIaN (North), VIaS (South) and VIb areas. The stock identity of herring west of the British Isles was reviewed by the EU-funded project WESTHER, which identified VIaN as an area where catches comprise a mixture of fish from Areas VIaN, VIaS, and VIIaN. ICES current advice is that herring components should be managed separately to afford maximum protection, but a study group will be convened in 2008 (SGHERWAY) to evaluate the WESTHER recommendations.

There are many hypotheses as to the cause of the irregular cycles shown in the productivity of herring stocks (weights-at-age and recruitment), but in most cases it is thought that the environment plays a key role (through prey, predation and transport). The VIaN herring stock has shown a marked decline in productivity during the late 1970s and has remained at a low level since then. ICES identifies that the VIaN stock is currently fluctuating at low levels and is being exploited above $F_{m s y}$.

Historically, the stock in this area has been affected by three fisheries:
A Scottish domestic pair trawl fleet and the North Irish fleet operated in shallower, coastal areas, principally fishing in the Minches and around the Island of Barra in the South where younger herring are encountered. This fleet has reduced in the last years.

The Scottish single-boat trawl and purse-seine fleets, with refrigerated seawater tanks, targeting herring mostly in the northern North Sea, but also operating in the northern part of VIaN. This fleet now operates mostly with trawls but many vessels can deploy either gear.

An international freezer-trawler fishery has historically operated in deeper water near the shelf edge where older fish are distributed. These vessels are mainly registered in the Netherlands, Germany, France, and England but most are Dutch owned.

In recent years the age structure of the catch of these last two fleets has become more similar.

In addition to being a valuable protein resource for humans, herring represent an important prey item for many predators including cod and other large gadoids, dogfish and sharks, marine mammals and sea birds. Because the trophic importance of herring puts its stocks under immense pressure from constant exploitation, it is important that management takes into account all anthropogenic, environmental and biological variables.

## H. 2 Terminology

The WG uses "rings" rather than "age" or "winter rings" throughout the report to denominate the age of herring, with the intention to avoid confusion. It should be observed that, for autumn spawning stocks, there is a difference of one year between "age" and "rings". HAWG in 1992 (ICES 1992/Assess:11) stated that:
"The convention of defining herring age rings instead of years was introduced in various ICES working groups around 1970. The main argument to do so was the uncertainty about the racial identity of the herring in some areas. A herring with one winter ring is classified as 2-years-old if it is an autumn spawner, and one-year-old if it is a spring spawner. Recording the age of the herring in rings instead of in years allowed scientists to postpone the decision on year of birth until a later date when they might have obtained more information on the racial identity of the herring.

The use of winter rings in ICES working groups has introduced a certain amount of confusion
and errors. In specifying the age of the herring, people always have to state explicitly whether they are talking about rings or years, and whether the herring are autumn- or spring spawners. These details tend to get lost in working group reports, which can make these reports confusing for outsiders, and even for herring experts themselves. As the age of all other fish species (and of herring in other parts of the world) is expressed in years, one could question the justification of treating West-European herring in a special way. Especially with the present trend towards multispecies assessment and integration of ICES working groups, there might be a case for a uniform system of age definition throughout all ICES working groups.

However, the change from rings to years would create a number of practical problems. Data files in national laboratories and at ICES would have to be adapted, which would involve extra costs and manpower. People that had not been aware of the change might be confused when comparing new data with data from old working group reports. Finally, in some areas (notably Division IIIa), the distinction between spring- and autumn spawners is still hard to make, and scientists preferred to continue using rings instead of years.
The Working Group discussed at length the various consequences of a change from rings to years. The majority of the Group felt that the advantages of such a change did not outweigh the disadvantages, and it was decided to stick to the present system for the time being. "

The text table below gives an example for the correlation between age, rings and year class for the different spawning types in late 2002:

| Year class (autumn spawners) | $\mathbf{2 0 0 1 / 2 0 0 2}$ | $\mathbf{2 0 0 0} / \mathbf{2 0 0 1}$ | $\mathbf{1 9 9 9} / \mathbf{2 0 0 0}$ | $\mathbf{1 9 9 8 / 1 9 9 9}$ |
| :--- | :--- | :--- | :--- | :--- |
| Rings | 0 | 1 | 2 | 3 |
| Age (autumn spawners) | 1 | 2 | 3 | 4 |
| Year class (spring spawners) | 2002 | 2001 | 2000 | 1999 |
| Rings | 0 | 1 | 2 | 3 |
| Age (spring spawners) | 0 | 1 | 2 | 3 |

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# Annex 7 - Stock Annex Herring in Division Vla South and VIIbc 

\author{

Quality Handbook ANNEX: Herring VIaS and VIIb, c <br> Stock specific documentation of standard assessment procedures used by ICES. <br> \begin{tabular}{ll}
Stock: \& Herring in VIaS and VIIb, c <br>

Working Group: \& | Herring Assessment Working Group for the area south of |
| :--- |
|  |
|  |
| $62^{0} \mathrm{~N}$ | <br>

Date: \& March 2008 <br>
Authors: \& Afra Egan and Maurice Clarke
\end{tabular}

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## A. General

The herring to the northwest of Ireland comprise both autumn and winter/spring spawning components. The age distribution of the catch and vertebral counts were used to distinguish these components (Bracken, 1964, Kennedy, 1970). Spawning takes place from September until March and may continue until April (Molloy and Kelly, 2000). Spawning in VIIb has traditionally taken place in the autumn and in VIaS, later in the autumn and in the winter.

For the purpose of stock assessment and management, these areas have been separated from VIaN since 1982 and are split at $56^{\circ} \mathrm{N}$. This split is based on work carried out by working groups in the late 1970s and early 1980s which found that the stocks exploited off the west coast of Scotland were biologically different from those off the north coast of Ireland. A second new assessment area was also recommended by the 1981 Working Group (ICES CM 1981). The Irish landings were taken mainly in the southern part of VIa and in VIIb, c. These catches were found to be biologically very similar with respect to age composition and spawning. It was decided at the 1981 working group to combine the areas and conduct a joint assessment (Molloy, 2006).

A herring tagging experiment was carried out in 1992 in order to investigate the movements and annual migrations of herring around the Irish Coast. 20,000 herring were tagged in total with 10,000 of these off the west coast. Some fish moved northwards and were recaptured along the north coast between July and February, in the main fishing areas. $90 \%$ of the fish tagged along the west coast were recovered from the Donegal Bay area. The maturity stages of the recaptured fish, suggests that the fish were migrating inshore towards spawning grounds (Molloy, et al 1993). There were no returns from north of Donegal although it is possible that there may not have been much fishing activity in the area at this time (Molloy and Kelly, 2000).

## Assessment and biology

A study group on herring assessment and biology in the Irish Sea and adjacent areas met in 1994 (ICES, 1994). This meeting highlighted the problems associated with the assessment of herring stocks around Ireland. This group recommended that the boundary line separating this stock from the herring stock of VIaS and VIIb be moved southwards from latitude $52^{\circ} 30^{\prime} \mathrm{N}$ to $52^{\circ} 00^{\prime} \mathrm{N}$ (ICES, 1994). A Schematic presentation of the life cycle of herring to the west and northwest of Ireland is shown in Figure
A.1. The spawning, nursery and feeding grounds are shown as well as the direction of larval drift and migration.

## WESTHER

WESTHER was an EU-funded project, to review, the stock identity of herring west of the British Isles. A number of factors were examined including.

Morphometrics and meristic characteristics
Internal parasites
Otolith microstructure and microchemistry
Genetics
Results from this project identified distinct spawning grounds and spawning components. It was recommended that the stocks to the west of the British Isles should be managed as two stocks, the Malin Shelf stock and the Celtic Sea stock. Management plans should be fleet and area based in order to prevent the local depletion of any population unit in the areas (WESTHER, Q5RS-2002-01056). Further work on the management of these stocks will be conducted by SGHERWAY which are due to meet in late 2008.

## A.2. Fishery

## Development of this fishery

In the early 1900s the main herring fisheries in Ireland were located off the Donegal coast. Donegal matje herring was important in supplying the German markets. Herring fisheries, which took place every spring and summer off the coast of Donegal, have been under scientific observation since 1921, with very little scientific work carried out prior to this. The fishing grounds were well known and were located between ten and forty miles offshore. Fishing during this time was split into three well defined time periods.

1) December/January

2 ) May (main fishing took place)
3 ) September/October
During the 1930s many of the major herring markets disappeared (Molloy, 1995). In contrast to the rapid expansion experienced in the Celtic Sea the revival of the northwest fishery occurred at a slower pace (Molloy, 2006). The revival first became evident in the 1950s when many Scottish ring netters took part in this fishery with many of the Irish boats also using this gear. Then several boats changed to pelagic midwater trawls. The herring fleet continued to expand throughout the 1960s with many skippers becoming experts in pelagic pair trawling (Molloy, 2006).

In the 1970s and 1980s the autumn spawners became more significant and accounted for the majority of the landings. Galway and Rossaveal gained increasing importance as herring ports in the 1970s. In the 1974/75 season landings decreased dramatically and it was the first indication that the stock might have started to decline. The North Sea stock was already in decline and many Dutch boats were fishing off the Irish west coast. TACs were reduced and the stock continued to decline. In 1978 it was advised that the fishery be closed (Molloy, 2006). This closure lasted until 1981 and was re-
opened with new management units. VIaS and VIIb, c were joined and were assessed separately from VIaN.

In recent years the northern grounds have regained importance with catch also coming from the west coast close to the VIa boundary line (ICES, 2005). Very little fishing now takes place on previously important grounds in Galway Bay and along the Mayo coast (Molloy and Kelly, 2000).

Since the late 1970s considerable changes have taken place in the type of pelagic fishing carried out by Irish boats off the North West Coast, with directed herring fishing having been largely replaced by mackerel fishing (Breslin, 1998).

## Recent

The TAC is taken mainly by Ireland, which has over $90 \%$ of the quota. In recent years, only Ireland has exploited herring in this area. The fishery is concentrated in quarters one and four. Landings have decreased markedly from about $44,000 \mathrm{t}$ in 1990 to around 13,800 t in 2004 . Working group catches in the last two years have decreased over $17,000 \mathrm{t}$ in 2007 to over 13,000 in 2008. Total catch over the complete time series are shown in the Figure 3. The number of boats participating in this fishery remained constant for a number of years at around 30 vessels. Increases were seen in the last two years with 62 vessels landing northwest herring in 2008. The number of vessels engaged in fishing for herring depends very much on the availability of mackerel or horse mackerel. Many of the larger vessels target these species primarily.

The majority of the landings in recent years are taken in quarters one and four with small quantities landed in quarter three. The main age groups are $2,3,4$ and 5 with older age groups accounting for small proportions of the catch. The proportions of older age groups have been decreasing over the last number of years.

## A.3. Ecosystem aspects

Divisions VIaS and VIIb, c are located to the North West and west of Ireland respectively. This area is limited to the southwest by the Rockall Trough, where the transition between the Porcupine Bank and the trough is a steep and rocky slope with reefs of deepwater corals; further north, the slope of the Rockall Trough is closer to the coast line; west of the shelf break is the Rockall Plateau with depths of less than 200 m . The shelf area consists of mixed substrates, with soft sediments (sand and mud) in the west and more rocky, pinnacle areas to the east. The area has several seamounts: the Rosemary Bank, the Anton Dohrn sea mount and the Hebrides, which have soft sediments on top and rocky slopes (ICES, 2007b).

The shelf circulation is influenced by the poleward flowing 'slope current', which persists throughout the year north of the Porcupine Bank, but is stronger in the summer. A schematic representation of the oceanographic conditions in this area is presented in Figure A.2. Over the Rockall plateau, domes of cold water are associated with retentive circulation. Thermal stratification and tidal mixing generate a northwards running coastal current known as the Irish coastal current which runs northwards along the west coast (ICES, 2007). The main oceanographic features in these areas are the Islay and the Irish Shelf fronts. The waters to the west of Ireland are separated by the Irish shelf front. This front causes turbulence and this may bring nutrients from deep waters to the surface. This promotes the growth of phytoplankton and dinoflagellates where there is increased stratification. Associated with this is increased growth of zooplankton and aggregations of fish. The Islay front persists throughout the winter due to the stratification of water masses of different salinities
(ICES, 2006). The ability to quantify any variability in frontal location and strength is an important element in understanding fisheries recruitment (Nolan and Lyons, 2006).

In the North, most of the continental shelf is exposed to prevailing sothwesterly winds and saline oceanic waters cross the shelf edge between Malin head off the north coast of Ireland and Barra head in the Outer Hebrides. The Irish shelf current flows northwards and then eastwards along the north coast of Ireland (Reid et al, 2003). Freshwater discharges from rivers such as the Shannon and Corrib interact with the Eastern North Atlantic water on the Irish shelf front to produce the observed circulation pattern (ICES, 2006).

Sea surface temperature data have been collected from Malin head on the North coast of Ireland since 1958. During periods of low winter temperatures, there is less pronounced heating during the summer. This can be seen in 1963, 1978 and 1985-1986. During these years there were also stormy conditions. This is concurrent with the lower winter temperatures (ICES, 2007). There is considerable variability over the complete time series. A definite trend can be identified from the early 1990s. Since 1990 sea surface temperatures measured at stations along the northwest coast of Ireland have displayed a sustained increasing trend, with winter temperatures $>6^{\circ}$ and higher summer temperatures during the same period (Figure A.4), (Nolan and Lyons, 2006).

Environmental conditions can cause significant fluctuations in abundance in a variety of marine species including fish. A study conducted in 1980 found that west coast herring catches showed strong correlations with temperature and salinity at a constant lag of three or four years. Oceanographic variation associated with temperature and salinity fluctuations appears to affect herring in the first year of life, probably during the winter larval drift (Grainger 1980a).

Productivity in this region is reasonably high on the shelf but drops rapidly west of the shelf break. This area is important for many pelagic fish species. The shelf edge is a spawning area for mackerel Scomber scombrus and blue whiting Micromesistius potassou. Historically, there were important commercial fisheries for many demersals species also. On the shelf, the main resident pelagic species is herring Clupea harengus (ICES, 2007b). Preliminary examination of productivity shows that overall productivity in this area is currently lower than it was in the 1980s. Further information on this can be found in the HAWG report 2007 (ICES CM 2007).

Larvae that were spawned on the west and northwest coast follow a northwards drift. Larvae spawned further north off the Donegal coast were found to drift towards the Scottish west coast (Grainger and McArdle, 1985; Molloy and Barnwall, 1988) Studies have shown that the maximum larval depth is below the surface between $5-15 \mathrm{~m}$ and there has been no evidence of diel migration, or variation in the distribution of different larval size categories (Grainger 1980b). Galway Bay and Donegal Bay, several inshore lochs and also Stanton Bank, an offshore area northwest of the Irish north coast are important nursery areas (ICES, 1994; Anon., 2000).

The spawning grounds for herring along the northwest coast are located in inshore areas close to the coast. These spawning grounds may contain one or more spawning beds on which herring deposit their eggs. The timing of spawning is not the same on each spawning ground. Spawning grounds tend to be vulnerable to anthropogenic influences such as dredging and sand and gravel extraction.

## Discards

The main market for Irish herring in the late 1980s and early 1990s was the Japanese roe market. The development of this market coincided with a decline in a number of other herring markets. It was therefore only favourable to catch roe herring, whose ovaries are just at the point of spawning. This led to discarding of non roe herring due to the lack of a suitable market. The roe market is no longer the main market for Irish herring. It is not known what the level of discarding is in this stock area and if it is a problem in this fishery.

## By Catch

Overall there is a paucity of data relating to by catch and discarding in this area. Interactions between cetaceans and fishing vessels have not been well documented and therefore no information is available. It is not possible therefore to make assumptions regarding implications for the marine ecosystem in area VIaS and VIIb, c.

## B. Data

## B.1. Commercial Catch

The commercial catch data are provided by national laboratories belonging to the nations that have quota for this stock. In recent years, only Ireland has been catching herring in this area, and the data are derived entirely from Irish sampling. Sampling is performed as part of commitments under the EU Council Regulation 1639/2001.

Commercial catch at age data are submitted in Exchange sheet v 1.6.4. These data are usually processed using SALLOCL (Patterson, 1998b). However, since only one country participates in this fishery this system is not required. Ireland acts as stock coordinator for this stock.

## InterCatch

Since 2007, InterCatch, which is a web-based system for handling fish stock assessment data was used. National fish stock catches are imported into InterCatch. Stock coordinators then allocate sampled catches to unsampled catches, aggregate them to stock level and download the output. The InterCatch stock output can then be used as input for the assessment models. It is envisaged that this system will replace SALLOCL and other previously used systems.

## Reallocation of Catches

Since 2007, landings data were revised with respect to reallocation of catches between area VIaS and VIaN, for the years 2000-2005. Before 2000, a comprehensive reallocation was used. For 2000-2005, various procedures were used. These attempted to deal with the increasing Irish catches along the $56^{\circ}$ line and opportunistic Irish catches of herring in VIaN during the $4^{\text {th }}$ and $1^{\text {st }}$ quarter mackerel fishery. In some years some catches were reallocated, while in others no reallocations were made. In 2007, it was considered that the most correct procedure was that used before 2000. Therefore a retrospective reallocation has been conducted. It does not adequately consider the Irish herring catches in VIaN, nor does the reallocation consider fishing along the $56^{\circ}$ line. However, in the absence of better information on Irish directed herring fishing in VIaN, this procedure provides the best possible method.

## B.2. Biological

## Sampling Protocol

Landings data are available for this area from 1970. Data on catch numbers at age, mean weights at age and mean lengths at age are derived from Irish data. Sampling is conducted by area and by quarter. Landings from this fishery, at present, are mainly into the port of Killybegs with lesser amounts landed into Rossaveal. Irish samples are collected from these commercial landings. Length frequency and age data is collected by ICES division by quarter. The length frequency data is added together for each division and quarter and raised to the landings for that area and quarter. The sample weight is divided into the catch weight to get the raising factor. The sum of the length frequencies per quarter is multiplied by the raising factor. An age length key is applied to this data and catch numbers at age calculated.

## Age Reading Protocol

Northwest herring are currently aged using otoliths and are read using a stereoscopic microscope, with reflected light. The minimum level of magnification (15x) is used initially. It is then increased to resolve the features of the otolith. Herring otoliths are generally read in the magnification range of $20 x-25 x$. The patterns of opaque (summer) and translucent (winter) zones are viewed. The winter (translucent) ring at the otolith edge is counted only in otoliths from fish caught after the $1^{\text {st }}$ January. The first winter ring that is counted is that which corresponds to the second "birth date" of the fish. Therefore a fish of 2 winter rings is a 3 year old. This convention applies to all ICES herring stocks with autumn spawning (Lynch, in prep).

## Age composition in the catch

Scales were used in the past for ageing and on average 4 and 5 ringers counted for $46 \%$ of the total catch. In 1929 however strong year classes were evident with 4 and 5 ringers making up $85 \%$ of the total (Farran, 1928). Currently the catch is mainly composed of ages $2,3,4$ and 5 ringers. In recent years there have been decreasing proportions of older fish in the catch. This stock is different from the Celtic Sea in that there is no recruitment failure and the Northwest stock is less reliant on incoming recruitment. The decrease in the proportions of older ages can be seen in Figure B.1.

## Precision Estimates

The precision estimates on 2006 ageing data were worked up using a bootstrap technique. The results of the method found that the relative error is below $20 \%$ over the age range $2-6 \mathrm{wr}$. At older ages, estimates of NW herring show higher CVs which is likely to be due to the relative paucity in the catch.

## Mean Weights

Mean weights in the stock (West) are calculated using samples taken from Q1 and Q4. A mean weight at age is then calculated. Mean weights in the catch (Weca) are calculated using samples from all quarters of the fishery and a mean weight per age derived.

## Trends in mean weights over time

The mean weights in the catch display quite a stable pattern over the time series, although variable weights are only available from the early 1980s. Younger ages (1-6 ring) show an overall downward trend with more fluctuations evident in older ages ( $7-9$ ring). The mean weights in the stock at spawning time have been calculated from Irish samples taken during the main spawning period and show similar patterns to the mean weight in the catch.

## Maturity ogive

A maturity ogive has been produced from the 2007 acoustic survey shows that $58 \%$ are mature at 1 -ring, $99 \%$ at 2 -ring and $100 \%$ mature at 3 -ring. The maturity ogive used in the assessment considers 1-ringers to be all immature and all subsequent age groups as fully mature.

## Log Catch Ratios

The log catch ratios ( $\ln \mathrm{C}_{\mathrm{a}, \mathrm{y}} / \mathrm{C}_{a+1, y+1}$ ) are presented below and are smoothed with a 4year running average to show the main trends (Figure B.2). Data for 1-ringers are noisy because this group is not fully selected by the fishery. The data for older fish are also noisy, particularly in later years, reflecting their relative paucity in the catches and suggest high variability in the exploitation rates of these age groups. These show an upward trend for all fully recruited year classes since the mid nineties. Overall, the catch data show a diminishing range of ages in the catches and older fish are at their lowest levels in the time series.

## Catch Curves

Cohort catch curves, were constructed for each year class in the catch at age data (Figure B.3). These catch curves show signals in total mortality over the time series. Low mortality seems evident on the very large 1981,985 and 1988 year classes. These represent three of the biggest year classes recruited to this fishery. Increasing mortality can be seen from 1990 on, whilst the 1970s cohorts show lower Z.

## B.3. Surveys

## Acoustic Surveys

Acoustic surveys have been carried out in this area since 1994. The timing of these surveys has changed over this period. Initially the surveys were undertaken in the summer in order to coincide with international herring surveys and with the summer feeding period of this stock. In 1997, a research vessel was not available and the survey was not carried out. From 1998-2001 surveys were undertaken in October in order to survey the autumn spawning component. This was changed in 2002 with surveys carried out in January targeting the winter spawning components of this stock.

Since 2004 the surveys have been carried out on the R.V. Celtic Explorer. A parallel transect design was adopted with transects running perpendicular to the coastline and extending up to 54 nmi (nautical miles) offshore. Transect spacing was set at 2 nmi throughout the survey. In bays a single zigzag transect approach was used to optimise coverage. The survey area was divided into strata based on the timing of spawning in each area. The first strata to be covered was chosen in order to contain
the earliest spawning components of the stock. The second strata is characterised as containing a mixture of early and mid spawning stock components. The third strata covered the area where the latest spawning is known to occur. Strata were subdivided in order to concentrate on known spawning grounds.

The acoustic data were collected using the Simrad ER60 scientific echosounder. The Simrad ES-38B ( 38 KHz ) split-beam transducer is mounted within the vessels drop keel and lowered to the working depth of 3.3 m below the vessels hull or 8.8 m below the sea surface.

Acoustic data analysis was carried out using Sonar data's Echoview® (V 3.2) post processing software and was backed up every 24 hrs. Partitioning of data was viewed and agreed upon by 2 scientists experienced in viewing echograms. Where no directed trawling had taken place, biological data from the nearest neighbour was used to determine the size classification of the echotrace.

The following TS/length relationships were used to analyse the data.
Herring $\quad \mathrm{TS}=20 \log \mathrm{~L}-71.2 \mathrm{~dB}$ per individual
( $\mathrm{L}=$ length in cm )
Sprat $\quad \mathrm{TS}=20 \log \mathrm{~L}-71.2 \mathrm{~dB}$ per individual
( $\mathrm{L}=$ length in cm )
Mackerel
$\mathrm{TS}=20 \log \mathrm{~L}-84.9 \mathrm{~dB}$ per individual
( $\mathrm{L}=$ length in cm )
Horse mackerelTS $=20 \log \mathrm{~L}-67.5 \mathrm{~dB}$ per individual $(\mathrm{L}=$
length in cm )
The winter acoustic survey time series was split and ran from 1999-2003 and 20042007 because of the timing. Earlier survey series were carried out in Q4 and the more recent surveys were in Q1. The acoustic survey time series is shown in the text table below. A problem with the winter acoustic survey series has been synchronising the survey with the peak spawning event to ensure containment of the stock. The winter surveys that were carried out from 2004-2007 varied sharply in age profile and biomass estimates, and was not considered reliable. Bad weather often affected the survey as it took place in January. Also it was recognised that synoptic coverage of a stock that spawns over a period from October to February in an area spanning all of Divisions VIaS and VIIb cannot be achieved with a winter survey. Thus the series was discontinued in 2007. The review group of the 2007 assessment highlighted that although there is an acoustic abundance estimate, the historical series is too short to consider it as a tuning survey in an analytical assessment.

Acoustic surveys have been conducted in this area since 1999. In the mid 1990s, surveys were undertaken in summer. The timing changed in 1999 with the surveys being carried out in the winter (Table 6.3.1). Table 6.3 .2 shows acoustic abundance at age and biomass estimates from all surveys conducted in this area, since 1994. The WESTHER project recommended that the survey effort along the Malin shelf area (including VIaN, VIaS, VIIb,c, Clyde and Irish Sea) should be increased or diverted to a combined survey on non-spawning herring. In 2008 PGHERS (CM 2008/LRC:01) discussed the possibility of conducting synoptic summer surveys on the Malin shelf.

The WESTHER project recommended that the survey effort along the Malin shelf area (including VIaN, VIaS, VIIb,c, Clyde and Irish Sea) should be increased or diverted to a combined survey on non-spawning herring. In 2008 PGHERS (CM 2008/LRC:01) discussed the possibility of conducting synoptic summer surveys on the Malin shelf.

In 2008, the Irish survey of VIaS, VIIb, c was conducted in July with effort concentrating on summer feeding aggregations.

## Larval Surveys

Assessment of this stock was largely based on the results of larval surveys in the 1980s. Herring Larval surveys were first carried out on this stock, by Ireland, in 1981 and continued until 1986. Prior to this the surveys were carried out by the Scottish but only had limited coverage of the assessment area. The survey grid consisted of sampling stations about 18 km apart. A gulf III plankton sampler with $275 \mu \mathrm{~m}$ mesh was towed at each station. The samples collected were preserved in $4 \%$ formalin. Herring larvae were identified and measured. Only larvae of less than 10 mm were used for the assessment. The number of larvae below each square meter was calculated and then multiplied by the area of the sea at each station (Grainger and McArdle, 1981). These surveys did not produce a satisfactory index of stock size because of two very low values in 1984 and 1985 (Molloy, 1989). However these surveys did provide valuable information on the distribution of very small larvae and on the location of the spawning grounds (Molloy and Kelly, 2000).

## Ground Fish Survey

The IGFS is part of the western IBTS survey and has been carried out on the RV Celtic Explorer since 2003. The gear used on the survey is a GOV 36/47 demersal trawl with a 20 mm cod end liner to retain juvenile and small fish, including small herring. This survey has been conducted since the early 1990s but is of little utility as a herring recruit index, because the gear, timing and survey vessel changed throughout. Once a sufficient time series becomes available it will be investigated as a possible tuning fleet. The Scottish groundfish survey, which has some coverage of VIaS will also be investigated as an additional tuning fleet.

## B.4. Commercial CPUE

Research surveys were not started in Ireland until the mid 1960s and in the absence of this information commercial catch per unit effort (CPUE) data was used as an index of stock size. It is known that CPUE data may not give an accurate index of stock size due to the shoaling nature of pelagic stocks. Fish can aggregate in dense shoals in a small area and CPUE may remain high even though the stock size is low. However the CPUE data collected in the 1960s and 1970s did provide an index of changes that were occurring in the fisheries around Ireland. F was calculated for the Northwest herring stock using this data during this time and showed an increasing trend in F. This CPUE data was used to show the dramatic decline that took place in this stock in the 1970s (Molloy, 2006).

## C. Historical Stock Development

## Time periods in the fishery

This fishery peaked in the late 1980s, largely as a result of two strong year classes in 1981 and 1985. This corresponded to the highest SSB and a medium level of F. In the late 1980s changes also took place with regard to the location and timing of the fishery. The North and West coast fisheries in December and January were now the most important with smaller amounts taken during the autumn fishery (Molloy, 2006). Since then there has been a downward trend in SSB and recruitment with no evi-
dence of strong year classes entering the fishery. Mean F has been fluctuating but is though to be at a high level.

Spawning stock size peaked in 1988 and has followed a steady decline since then. Landings have drastically fallen since 1999 (ICES, 2004). Long term changes in the spawning component have occurred in the area and time of spawning. In 1920-1930s there was a north coast fishery that spawned in the North in spring and an autumn fishery that spawned in the west of Donegal. Sligo and Galway had no important fishery. In the '40-50 herring all over Ireland declined and the recovery in the 1960s occurred mainly in Mayo, Sligo and Galway as autumn spawners. Recently there has been a shift to the northern fishery, while little fishing occurs on the west coast of Ireland. The northwest herring fishery was based on hard (stage V) herring but towards the late 1980s the focus shifted to spawning herring.

## Assessment

In 1930, Farran made his first attempt to quantify the abundance of the herring stock in this area. In the 1930s many of the previous herring markets disappeared and there was widescale discarding of herring along the Donegal coast. It is thought that during this time that the herring population was at a very low level (Molloy, 1995).

## Recent Assessments

In recent years the model used for this stock was a separable VPA. This was used to screen over three terminal fishing mortalities, $0.2,0.4$ and 0.6 . In 2009 terminal $F$ of 0.5 was also examined. This was achieved using the Lowestoft VPA software (Darby and Flatman, 1994). Reference age for calculation of fishing mortality was 3-6 and terminal selection was fixed at 1, relative to age 3 winter rings. ICA was used in exploratory assessments with the acoustic surveys as a tuning fleet.

## Model used: ICA and VPA

No final assessment has been accepted for this stock by the working group. However several scenarios are run, screening over a range of terminal F's ( $0.2,0.4,0.5$ and 0.6 ). In 2006 and 2007 exploratory runs using the ICA model (Patterson, 1998) were performed. In the absence of a sufficient time series in this area the use of the ICA model has discontinued.

Software used: VPA
A separable VPA is used to track the historic development of this stock.
Software used: Lowestoft VPA Package (Darby and Flatman, 1994).
VPA SETTINGS
Reference Age $=3$
Selection in the terminal year $=1.0$
Terminal F $=0.2,0.4,0.5,0.6$
1 Ringers: downweighted to 0.1
Reference ages for calculation of Mean $\mathrm{F}=3-6$
Software used: ICA (exploratory runs in 2006 and 2007 only)
Model Options chosen:
Separable constraint over the last 6 years (weighting = 1.0 for each year)
Reference ages: 3
Constant selection pattern model
Selectivity on oldest age: 1.0
First age for calculation of mean F: 3
Last age for calculation of mean F: 6
Weighting on 1 ringers: 0.01 Other age classes: 1.0
Lowest feasible F: 0.05
Highest feasible F: 2.0
Ages for acoustic abundance estimates: 3-4
Plus group: 9

Input data types and characteristics:

| TYPE | NAME | YEAR <br> RANGE | AGE <br> RANGE | VARIABLE FROM YEAR TO <br> YEAR <br> YES/NO |
| :--- | :--- | :--- | :--- | :--- |
| Caton | Catch in tonnes | $1970-$ <br> 2008 | $1-9+$ | Yes |
| Canum | Catch at age in numbers | $1970-$ <br> 2008 | $1-9+$ | Yes |
| Weca | Weight at age in the commercial <br> catch | $1970-$ <br> 2008 | $1-9+$ | Yes |
| West | Weight at age of the spawning <br> stock at spawning time. | $1970-$ <br> 2008 | $1-9+$ | Yes |
| Mprop | Proportion of natural mortality <br> before spawning | $1970-$ <br> 2008 | $1-9+$ | No |
| Fprop | Proportion of fishing mortality <br> before spawning | $1970-$ <br> 2008 | $1-9+$ | No |
| Matprop | Proportion mature at age | $1970-$ <br> 2008 | $1-9+$ | No |
| Natmor | Natural mortality | $1970-$ <br> 2008 | $1-9+$ | No |

Tuning data:

| TYPE | NAME | YEAR RANGE | AGE RANGE |
| :--- | :--- | :--- | :--- |
| Tuning fleet 1 | NWHAS | $1999-2003$ | $3-4$ |
| Tuning fleet 2 | NWHAS | $2004-2007$ | $3-4$ |

## D. Short-Term Projection

Due to the absence of information on recruitment and the uncertainty about the current stock size short term predictions have not been routinely carried out for this stock.

## E. Medium-Term Projections

Model Used: Multi Fleet Yield Per Recruit
Software Used: MFYPR Software
Yield-per-recruit analysis was carried out using MFYPR to provide yield-per-recruit plots for the data produced in the assessment. The values for $\mathbf{F}_{0.1}$ and $\mathbf{F}_{\text {med }}$ are 0.17 and 0.31. $\mathrm{F}_{\max }$ is undefined and this is consistent with many other pelagic species (ICES, 2006).

## F. Long-Term Projections

Not performed

## G. Biological Reference Points

In 2007 the technical basis for the selection of the precautionary reference points was examined based on methods used by SGPRP (ICES CM 2001). No alternative biomass and fishing mortality reference points are available. It is clear that recruitment does not show any clear dependence on the SSB and that apart from the very high year classes in the 1980s is showing a decline.

The SGPRP (ICES CM 2003) has reviewed the methodology for the calculation of biological reference points, and applying a segmented regression to the stock and recruit data from the 2002 HAWG assessments. This showed that the fit to the stock and recruit data for this stock was not significant. There was no well defined change point and there was no reason to refine the reference points at that time.

## Current reference points

$B_{p a}=81,000 t=$ the lowest reliable estimate of SSB
$B_{\text {llim }}=110,000 t=1.4 \times B_{\text {pa }}$
$\mathrm{F}_{\mathrm{pa}}=0.22=\mathrm{F}_{\text {med }}$ (1998)
$\mathrm{F}_{\text {lim }}=0.33$ = lowest observed F

## H: Other Issues

## H. 1 Biology of the species in the distribution area

The herring (Clupea harengus) is a widely distributed pelagic species in this area. This stock is comprised of different spawning components. Off the west coast the majority of the stock, are autumn spawners. Off the northwest coast distinct spawning units have also been identified. Autumn spawners, that spawn in the Donegal Bay area and winter/spring spawners, that spawn further north off the Donegal coast (Breslin, 1998). Autumn and winter spawners were distinguished by vertebral counts and timing of maturity.

Herring are benthic spawners and deposit their eggs on the sea bed usually on gravel or course sediments. The yolk sac larvae hatch and adopt a pelagic mode of life.

When referring to spawning locations the following terminology is used (Molloy, 2006)

- A spawning bed is the area over which the eggs are deposited
- A spawning ground consists of one or more spawning beds located in a small area.
- A spawning area is comprised of a number of spawning grounds in a larger area

Spawning grounds are typically located in high energy environments such as the mouth of large rivers and areas where the tidal currents are strong. Herring shoals return to the same spawning grounds each year (Molloy, 2006).

The spawning grounds for northwest herring are located in shallow waters close to the coast and are well known and well defined. Spawning begins in October and can continue until February. Fecundity is the number of eggs produced by the female and is proportional to the length of the fish (Molloy, 2006). Several studies were carried out in the early 1980s to analyse the fecundity of winter and autumn spawning components of the North West herring stock and considerable differences were found.

Donegal winter spawners produce significantly fewer eggs than autumn spawners. When compared to the Celtic Sea herring stock, Donegal herring have a higher fecundity and begin to spawn earlier (McArdle, 1983). A study conducted in the 1920s found that the eggs produced by winter/spring spawners were $25 \%$ bigger than those autumn spawners but were less numerous (Farran, 1938).

Herring produce benthic eggs that are adhered to the bottom substrate where they remain until the larvae hatch. The larvae are carried by the currents and drift towards the west coast of Scotland (Grainger and McArdle, 1985). Several important nursery grounds for juveniles have been identified in this area.

The larval phase is an important period in the herring life cycle. Larvae use their oil globule for food and to provide buoyancy. Their movements and survival are determined by favourable environmental conditions. Larvae originating from spawning grounds off the west coast are carried by currents to the northwest coast of Donegal and may even travel as far as Scotland (Molloy, 2006). Figure A. 1 shows a schematic presentation of the life cycle of Herring west and northwest of Ireland.

The juveniles tend to remain close inshore, in shallow waters for the first two years of their lives, in nursery areas. There are many of these nursery areas around the coast, for example St. Johns point in Donegal Bay. The minimum landing size for herring is 20 cm and therefore these juvenile herring are not caught by the fishery in the early stages of their life cycle (Molloy, 2006).

Changes in the growth rate of this stock can be seen over time. In the late 1980s a sudden and unexplained drop in mean weights was observed. This had an impact on the estimate of SSB and the advised TAC. The growth rate of this stock has never recovered to the levels before this decline (Molloy, 2006).

Adult herring are found offshore until spawning time, when they move inshore. Occasionally very large herring are found off the Irish coast. Theses herring appear off the north coast and are usually in a spawning or pre spawning condition (Molloy, 2006).

## H.2. Management and ACFM advice

## Local Management

Various management measures have been introduced to control the exploitation of this stock. From 1972-1978 TACs were set by NEAFC and covered all of area VIa. The TAC decreased rapidly and the stock was thought to be in decline. This continued until the fishery was closed in 1979 and 1980. During the closure because there was no analytical assessment of VIIb fishing was allowed to continue on a precautionary basis (ICES, 1994). When the fishery was reopened it was decided to split the area into VIaS and VIaN. Landings from this area increased due to the increased efficiency of the Irish vessels and the participation in this fishery by Dutch vessels (Anon, 2000).

Management measures were slowly introduced into this fishery with by-laws restricting fishing in certain areas off the coast in the early 1900s. This type of management continued until the 1930s when fishing was prohibited during April and May, in order to improve the quality of the herring being landed. In the 1970s management measured became more defined. Direct fishing of herring for fishmeal was banned. A minimum landing size of 20 cm was implemented and also minimum mesh sizes. TACs were introduced in order to control the amount of herring landing each year from each ICES area (Molloy, 1995).

The management of the fishery has improved in recent years and catches have been considerably reduced since 1999. In 2000 the Irish North West Pelagic Management Committee was established to deal with the management of this stock. The assessment period runs concurrently with the annual quota. Quotas are allocated on a fortnightly basis and there is some capacity to carry unused allocation into the following fortnight with overruns being deducted.
In 2000, the Irish North West Pelagic Management Committee was established to deal with the management of this stock. The committee has the following objectives:

- To rebuild this stock to above the $B_{p a}$ level of 110000 t .
- In the event of the stock remaining below this level, additional conservation measures will need to be implemented.
- In the longer term it is the policy of the committee to further rebuild the stock to the level at which it can sustain annual catches of around 25000 t .
- Implement a closed season from March to October.
- Regulate effort further through boat quotas allocated on a weekly basis in the open season.

This committee manages the whole fishery for this stock at present, given that Ireland currently accounts for the entire catch.

The current state of the stock is uncertain. Preliminary assessments suggest that SSB may be stable at a low level. The current level of SSB is uncertain but likely to be below $\mathrm{B}_{\mathrm{lim}}$. There is no evidence that large year classes have recruited to the stock in recent years. F appears to have increased concomitantly with increases in the catch. F is likely to be above $\mathrm{F}_{\mathrm{pa}}$ and also likely above $\mathrm{F}_{\text {lim }}$.

There is no explicit management plan for this stock. The local Irish management committee developed the objective to rebuild the stock to above $\mathrm{B}_{\mathrm{pa}}$ and to maintain catches of 25000 t per year. The implementation of the closed season from March to October has been successful in ensuring that the fishery mainly concentrates on the spawning component in this area. ICES have recommended that a rebuilding plan be put in place that will reduce catches. If no rebuilding plan is established, there should be no fishing. The rebuilding plan should be evaluated with respect to the precautionary approach.

## H. 4 Terminology

The WG uses "rings" rather than "age" or "winter rings" throughout the report to denominate the age of herring, with the intention to avoid confusion. It should be observed that, for autumn spawning stocks, there is a difference of one year between "age" and "rings". HAWG in 1992 (ICES 1992/Assess:11) stated that
"The convention of defining herring age rings instead of years was introduced in various ICES working groups around 1970. The main argument to do so was the uncertainty about the racial identity of the herring in some areas. A herring with one winter ring is classified as 2-years-old if it is an autumn spawner, and one-year-old if it is a spring spawner. Recording the age of the herring in rings instead of in years allowed scientists to postpone the decision on year of birth until a later date when they might have obtained more information on the racial identity of the herring.

The use of winter rings in ICES working groups has introduced a certain amount of confusion and errors. In specifying the age of the herring, people always have to state explicitly whether they are talking about rings or years, and whether the herring are autumn or spring spawners. These details tend to get lost in working group reports, which can make these reports confusion for outsiders, and even for herring experts themselves. As the age of all other fish species (and of herring in other parts of the world) is expressed in years, one could question the justification of treating WestEuropean herring in a special way. Especially with the present trend towards multispecies assessment and integration of ICES working groups, there might be a case for a uniform system of age definition throughout all ICES working groups.

However, the change from rings to years would create a number of practical problems. Data files in national laboratories and at ICES would have to be adapted, which would involve extra costs and manpower. People that had not been aware of the change might be confused when comparing new data with data from old working group reports. Finally, in some areas (notably Division IIIa), the distinction between spring and autumn spawners is still hard to make, and scientists preferred to continue using rings instead of years.

The Working Group discussed at length the various consequences of a change from rings to years. The majority of the Group felt that the advantages of such a change did not outweigh the disadvantages, and it was decided to stick to the present system for the time being."

The text table below gives an example for the correlation between age, rings and year class for the different spawning types in late 2002:

| YEAR CLASS (AUTUMN SPAWNERS) | $\mathbf{2 0 0 1 / 2 0 0 2}$ | $\mathbf{2 0 0 0 / 2 0 0 1}$ | $\mathbf{1 9 9 9 / 2 0 0 0}$ | $\mathbf{1 9 9 8 / 1 9 9 9}$ |
| :--- | :--- | :--- | :--- | :--- |
| Rings | 0 | 1 | 2 | 3 |
| Age (autumn spawners) | 1 | 2 | 3 | 4 |
| Year class (spring spawners) | 2002 | 2001 | 2000 | 1999 |
| Rings | 0 | 1 | 2 | 3 |
| Age (spring spawners) | 0 | 1 | 2 | 3 |

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Figure A. 1 Schematic presentation of the life cycle of Herring west and northwest of Ireland. Numbers represent locations mentioned in the text:1 - Dingle Peninsula, 2 - Shannon River, 3 Galway Bay, 4 - Mayo, 5 - Donegal Bay (ICES, 2005b, SGRESP)


Figure A. 2 Schematic presentation of prevailing oceanographic conditions in the west and northwest of Ireland. Fronts are 1.) the Islay front northeast of Ireland and 2.) the Irish shelf front to the west of the Celtic Sea, both fronts are a thermohaline fronts persisting throughout the year with an additional tidal mixing front developing near Islay during summer stratification. Residual currents are the Irish coastal current, a clockwise density current and the Atlantic shelf edge current. Circulation is mainly wind driven with prevailing south-easterly winds from October to May and density driven from May to October (ICES, 2005b, SGRESP).


Figure A.3: Total landings from VIaS, VIIb,c


Figure A.4: Sea surface temperature anomaly at Malin Head (1960-2005) (Nolan and Lyons, 2006)


Figure B.1: Mean Standardised Catch Numbers at Age

B.2: Log Catch Ratios with a four year running average


Figure B.3: Catch Curves by cohort

# Annex 8 - Stock Annex Irish Sea Herring VIIa (N) 

\author{

Quality Handbook ANNEX:_hawg-nirs <br> Stock specific documentation of standard assessment procedures used by ICES. <br> | Stock: | Irish Sea herring (VIIa(N) |
| :--- | :--- |
| Working Group | Herring Assessment Working Group (HAWG) |
| Date: | 25 March 2009 |
| Revised by | Steven Beggs |

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## A. General

## A.1. Stock definition

Herring spawning grounds in the Irish Sea are found in coastal waters to the west and north of the Isle of Man and on the Irish Coast at around $54^{\circ} \mathrm{N}$ (ICES, 1994; Dickey-Collas et al., 2001). Spawning takes place from September to November in both areas, occurring slightly later on average on the Irish Coast than off the Isle of Man. ICES Herring Assessment Working Groups from 19XX to 1983 used vertebral counts to separate catches into Manx and Mourne stocks associated with these spawning grounds. However, taking account of inaccuracies in this method and the results of biochemical analyses, the 1984 WG combined the data from the two components to provide a "more meaningful and accurate estimate of the total stock biomass in the N. Irish Sea." All subsequent assessments have treated the VIIa(N) data as coming from a single stock. During the 1970s, catches from the Manx component were about three times larger than those from the Mourne component. By the early 1980s, following the collapse of the stock, the catches were of similar magnitude. The fishery off the Mourne coast declined substantially in the 1990s then ceased, whilst acoustic and larva surveys in this period indicate that the spawning population in this area has been very small compared to the biomass off the Isle of Man.

The occurrence in the Irish Sea of juvenile herring from a winter-spring spawning stock has been recognized since the 1960s based on vertebral counts (ICES, 1994). More recently, Brophy and Danilowicz (2002) used otolith microstructure to show that nursery grounds in the western Irish Sea were generally dominated by winterspawned fish. Samples from the eastern Irish Sea were mainly autumn-spawned fish. Recaptures from 10,000 herring tagged off the SW of the Isle of Man in July 1991 occurred both on the Manx spawning grounds and along the Irish Coast with increasing proportions from the Celtic Sea in subsequent years (Molloy et al., 1993). The pattern of recaptures indicated a movement towards spawning grounds in the Celtic Sea as the fish matured.

A proportion of the Irish Sea herring stocks may occur to the north of the Irish Sea outside of the spawning period. This was indicated by the recapture on the Manx spawning grounds of 3-6 ring herring tagged during summer in the Firth of Clyde (Morrison and Bruce, 1981). Aggregations of post-spawning adult herring were detected along the west coast of England during an acoustic survey in December 1996 (Department of Agriculture and Rural Development for Northern Ireland, unpublished data), showing that a component of the stock may remain within the Irish Sea.

The results of WESTHER, a recent EU-funded programme aiming to elucidate stock structures of herring throughout the western seaboard of the British Isles have recently been published. Using a combination of morphometric measurements, otolith structure, genetics and parasite loads the conductivity of stocks within and beyond the Irish Sea have been examined. The results of this programme and existing knowledge are currently being evaluated at SGHERWAY in light of the future assessment and management of stocks to the western British Isles.

## A.2. Fishery

There have been three types of fishery on herring in the Irish Sea in the last 40 years:
i) Isle of Man- aimed at adult fish that spawn around the Isle of Man.
ii ) Mourne- aimed at adult fish that spawn off the Northern Irish eastern coast.
iii ) Mornington- a mixed industrial fishery that caught juveniles in the western Irish Sea.

The Mornington fishery started in 1969 and at its peak it caught 10,000 tonnes per year. It took place throughout the year. The fishery was closed due to management concerns in 1978 (ICES, 1994). In the 1970s the catch of fish from the Mourne fishery made up over a third of the total Irish Sea catch. The fishery was carried out by UK and Republic of Ireland vessels using trawls, seines and drift nets in the autumn. However the fishery declined and ceased in the early 1990s (ICES, 1994). The biomass of Mourne herring, determined from larval production estimates is now 2-4\% of the total Irish Sea stock (Dickey-Collas et al., 2001).

The main herring fishery in the Irish Sea has been on the fish that spawn in the vicinity of the Isle of Man. The fish are caught as they enter the North Channel, down the Scottish coast, and around the Isle of Man. Traditionally this fishery supplied the Manx Kipper Industry, which requires fish in June and July. However the fish appeared to spawn slightly later in the year in the 1990s and this lead to problems of supply for the Manx Kipper Industry. In 1998 the Kipper companies decided to buy in fish from other areas. Generally the fishery has occurred from June to November, but is highly dependent on the migratory behaviour of the herring.

The fishery has been prosecuted mainly by UK and Irish vessels. TACs were first introduced in 1972, and vessels from France, Netherlands and the USSR also reported catches from the Irish Sea during the 1970s before the closure of the fisheries from 1978 to 1981. By the 1990s only the fishery on the Manx fish remained, and by the late 1990s this was dominated by Northern Irish boats. The number of Northern Irish vessels landing herring declined from 24 in 1995-96 to 6-10 in 1997-99 and to 4 in 2000. Only two vessels operated in 2002 and 2003. However, total landings have remained relatively stable since the 1980s whilst the mean amount of fish landed per fishing trip has increased, reflecting the increase in average vessel size

## A.3. Ecosystem aspects

The main fish predators on herring in the Irish Sea include whiting (Merlangius merlangus), hake (Merluccius merluccius) and spurdog (Squalus acanthias). The size composition of herring in the stomach contents indicates that predation by whiting is mainly on 0-ring and 1-ring herring whilst adult hake and spurdogfish also eat older herring (Armstrong, 1979; Newton, 2000; Patterson, 1983). Sampling since the 1980s has shown cod (Gadus morhua), taken by both pelagic and demersal trawls in the Irish Sea, to be minor predators on herring. Small clupeids are an important source of food for piscivorous seabirds including gannets, guillemots and razorbills (ref...) which nest at several locations in and around the Irish Sea. Marine mammal predators in-
clude grey and harbour seals (ref.) and possibly pilot whales, which occur seasonally in areas where herring aggregate.

Whilst small juvenile herring occur throughout the coastal waters of the western and eastern Irish Sea, their distribution overlaps extensively with sprats (Sprattus sprat$t u s)$. The biomass of small herring has typically been less than $5 \%$ of the combined biomass of small clupeids estimated by acoustics (ICES, 2008 ACOM:02). However in recent years the proportions have increased in favour of small herring (ICES, 2009 ACOM:??).

There are irregular cycles in the productivity of herring stocks (weights-at-age and recruitment). There are many hypotheses as to the cause of these changes in productivity, but in most cases it is thought that the environment plays an important role (through transport, prey, and predation). Coincident periods of high and low production have been seen in the herring in VIaN and Irish Sea herring. Exploitation and management strategies must account for the likelihood of productivity changing. The Irish Sea herring stock has shown a marked decline in productivity during the late 70's and remained on a low level since then.

## Changes in Environment

There has been an increase in water temperatures in this area (ICES, 2006) which is likely to affect the distribution area of some fish species, and some changes of distribution have already been noted. Temperature increase is likely to affect stock recruitment of some species. In addition, the combined effects of over exploitation and environmental variability might lead to a higher risk of recruitment failure and decrease in productivity (ICES, 2007).

## B. Data

## B.1. Commercial catch

## National landings estimates

The current ICES assessment of Irish Sea herring extends back to 1961, and is based on landings only. ICES WG reports (ICES 1981, 1986 and 1991) highlight the occurrence of discarding and slippage of catches, which can occur in areas where adult and juvenile herring co-occur. Discarding has been practised on an increasing scale since 1980 (ICES, 1986). This increase is primarily related to the onset of slippage of catches that coincided with the cessation of the industrial fishery in early 1979 (ICES, 1980). As a result of sorting practices, slippage has led to marked changes in the age composition of the catch since 1979 and considerable change in the mean weights at age in the catch of the three youngest age groups (ICES 1981). Estimates of discarding were sporadically performed in the 1980s (ICES, 1981, 1982, 1985 and 1986), but there are no estimates of discarding or slippage of herring in the Irish Sea fisheries since 1986. Highly variable annual discard rates are evident from the 1980s surveys. For example, discards estimates of juvenile herring (0-group) for the Mourne stock taken in the 1981 Nephrops fishery was estimated at $1.9 \times 10^{6}$ of vessels landing in Northern Ireland, which amounts to approximately $20 \%$ of the Mourne fishery (ICES 1982). In 1982, at least $50 \%$ of 1-group herring caught were discarded at sea by vessels participating in the Isle of Man fishery (ICES, 1983). A more comprehensive survey programme to determine the rate of discarding in 1985 revealed discard estimates of $82 \%$ by numbers of 1 -ring fish, $30 \%$ of 2 -ring and $6 \%$ of 3 -ring fish, with the dominant age group in the landed catch being 3 ring (ICES, 1986). A similar survey in 1986, however, found the discarding of young fish fell to a very low level (ICES, 1987). The 1991 WG
discussed the discard problem in herring fisheries in general and suggested possible measures to reduce discarding. No quantitative estimates were given, but reports of fishermen suggesting discards of up to $50 \%$ of catch as a result of sorting practices by using sorting machines (ICES, 1991). The variation in discard rates since 1980, as a result of changes in discard practices, can probably be attributed to several changes in the management of the fishery. These include the availability of different fishing areas, the change to fortnightly catch quotas per boat (ICES, 1987) and level of TAC, where lower discard rates are observed with a higher TAC (ICES, 1989). The level of slippage is also related to the fishing season, since slippage is often at a high level in the early months (ICES, 1987). Due to the variable nature of discard estimates and the lack of a continuous data series, it has not been included in the annual catch at age estimates (with the exception of the 1983 assessment when the catch in numbers of 1ringers was doubled based on a $50 \%$ discard estimate of this age group).

Landings data for herring in Division $\mathrm{VIIa}(\mathrm{N})$ are generally collated from all participating countries providing official statistics to ICES, namely UK (England \& Wales, Northern Ireland, Scotland and the Isle of Man), Ireland, France, the Netherlands and what was formally the USSR. The data for the period 1971 to 2002 are reported in the various Herring Assessment Working Group Reports and are reproduced in Table 1. The official Statistics for Irish landings from VIIa have been processed to remove data from the Dunmore East fishery in area VIIa(S), and represent landings from VIIa(N) only.

Over the past three decades, the WG highlighted the under- or misreporting of catches as the major problem with regards to the accuracy of the landing data. Related to this are the problems of illegal landings during closed periods and paper landings. Area misreporting was also recognised (ICES, 1999), although a less prominent problem that is mostly corrected for.

The 1980 WG first identified the problem of misreporting of landings based on the results of a 3-year sampling programme, which was initiated after 1975 when herring were being landed in metric units at ports bordering the Irish Sea ( 1 unit $=100 \mathrm{~kg}$ nominal weight). The study showed the weight of a unit to be very variable, but was usually well in excess of 100 kg . An initial attempt to allow for misreporting using adjusted catches made very little difference to any of the values of fishing mortality (ICES, 1980). Subsequently, despite serious concerns about considerable underreporting being raised (ICES 1990, 1994, 2000 and 2001), the WG made no attempts to examination the extent of the problem. This uncertainty signifies no estimates of un-der-reporting and consequently no allowance for under-reporting of landings has been made. Considerable doubt was raised as to the accuracy of landing data over the period 1981-87 (ICES, 1994). However, after apparent re-examination all WG landing statistics are assumed to be accurate up to 1997 (ICES, 2000), but with no reliable estimates of landings from 1998-2000 (ICES, 2001). The WG acknowledged that poor quality landing data bring the catch in numbers at age data into question and hence the accuracy of any assessment using data from such periods (ICES, 1994).

In 2002 the ICES assessment was extended back to include data for 1961-1970 with the intention of showing the stock development prior to the large expansion in fishing effort and stock size in the early 1970s. This has now been extended further back to 1955. Landings data for this period were extracted from the UK fisheries data bases (England \& Wales, Scotland and Northern Ireland: Table 1, columns 8-10) and publications by Bowers and Brand (1973) for Isle of Man landings (column 11). Landings data for Ireland and France were not available.

To estimate the VIIa(N) herring landings for Ireland and France during 1955-1970, the NE Atlantic herring catches for each country were obtained from the FAO database (column 16). Using the ICES landings data for each country (column 17) the mean proportion of the VIIa(N) catch to the NE Atlantic catch during 1971 to 1981 was estimated (column 18). This was applied to the NE Atlantic catches from each country, for the period 1955 to 1970, to give an estimated landing for both France and Ireland (column 19). These landings were added to the known catches from the CEFAS database to give the total landings. The landings data (tonnes) used in the assessment are given in Table 1, column 14. It is anticipated that landings data for VIIa(N) for years prior to 1971 can be extracted from the Irish databases. However, the French landings will remain as estimates. As yet there has been no analysis of magnitude of errors in the old data. Need discussion on errors due to misreporting

## Catch at age data

Age classes in the ICES Canum file refer to numbers of winter rings in otoliths. As the Irish Sea stock comprises autumn spawners, $i$-ring fish taken in year $y$ will comprise fish in their $i_{\text {th }}$ year of life if caught prior to the spawning season and $(i+1)_{\text {th }}$ year if caught after the spawning period. An $i$-ring fish will belong to year-class $y$ - 2 . As spawning stock is estimated at spawning time (autumn), spawning stock and recruitment relationships require estimates of recruitment of $i$-ring fish in year $y$ and estimates of SSB in year $i-2$. The current assessment estimates recruitment as numbers of 1-ring fish.

The most recent description of sampling and raising methods for estimating catch at age of herring stocks is in ICES (1996). This includes sampling by UK(E\&W) and Ireland, but not UK(NI) and Isle of Man
$\mathrm{UK}(\mathrm{NI})$ : A random sample of $10-20 \mathrm{~kg}$ of herring is taken from each landing into the main landing port (Ardglass) by the NI Department of Agriculture and Rural Development. Samples are also collected from any catches landed into Londonderry. Prior to the 1990s, the samples were mostly processed fresh. During the 1990s, there was an increasing tendency for samples to be frozen for a period of weeks before processing. No corrections have been applied to weight measurements to allow for changes due to freezing and defrosting. The length frequency (total length) of each sample is recorded to the nearest 0.5 cm below. A sample of herring is then taken for biological analysis as follows: one fish per 0.5 cm length class, followed by a random sample to make the sample up to 50 fish.

Otoliths are removed from each fish, mounted in resin on a black slide and read by reflected light. Ages are assigned according to number of winter rings.

Length frequencies (LFDs) for VIIa(N) catches are aggregated by quarter. The weight of the aggregate LFD is calculated using a length-weight relationship derived from the biological samples. The LFD is then raised to the total quarterly landings of herring by the NI fleets. A quarterly age-length key, derived from commercial catch samples only, is applied to the raised LFD to give numbers at age and mean weight at age.

IOM: IOM sampling covers the period 1923 - 1997. Samples are collected from any landings into Peel, by staff of the Port Erin Marine Laboratory (Liverpool University). The sampling and raising procedures are the same as described for UK(NI) with the following exceptions: i) the weight of the aggregate quarterly LFD is obtained from the original sample weights rather than using a length-weight relationship, and ii) the biological samples are random rather than stratified by length. The 1993 ICES herring assessment WGs noted a potential under-estimation by one ring, of herring sampled
in the IOM. This was caused by a change in materials used for mounting otoliths and appears to have been a problem for ageing older herring in 1990-92. This was since rectified. However, the bias for the 1990-92 period has not yet been quantified and will be examined in the near future.

Ireland: Irish sampling of VIIa(N) herring covers the period 19xx - 2001. Some samples are from landings into NI but transported to factories in southern Ireland. Irish sampling schemes for herring in Div. VIa(S), VIIb, Celtic Sea and VIIj are described in ICES (1996). Methods for sampling catches in VIIa(N) are similar. The procedure is the same as described above for UK(NI) except that the biological samples are random rather than length stratified. ICES (1996) notes that a length-stratified scheme should be adopted to ensure proper coverage at the extremes of the LFDs.

Quality control of herring ageing has fallen under the remit of EU funded programmes EFAN and TACADAR, to which the laboratories sampling VIIa(N) herring contribute. An otolith exchange exercise was initiated in 2002 and is currently being completed.

## B.2. Biological

## Natural Mortality

Natural mortality (M) varies with age (expressed in number of winter rings) according to the following:

Rings M
$1 \quad 1$
20.3
30.2
$4+\quad 0.1$
Those values have been held constant from 1972 to date. Those values correspond to estimates for North Sea herring based on recommendations by the Multi-species WG (Anon. 1987a). which were applied to adjacent areas (Anon. 1987b).

## Maturity at age

Combined, year-specific maturity ogives were used in the 2003 Assessment (ICES 2003). The way those values were derived is documented on Dickey-Collas et al. (2003). Prior to 2003 annually invariant estimates of the proportion of fish mature by age were used. Those were based on estimates from the 1970s (ICES, 1994). The use of the variable maturity ogive in 2003 did not change greatly the perception of the stock state (Dickey-Collas et al., op cit). Due to inconsistencies in the maturity data collected in 2003, the WG used a mean maturity ogive for the preceding nine years for 2003. The rationale for the 9 years was that there appeared to be a shift in the maturity ogive around 1993. After 2003 all weights and maturity-at-age data were based on corresponding annual biological samples.

SSB in September is estimated in the assessment. The survey larvae estimate is used as a relative index of SSB. The proportions of M and F before spawning are held constant over time in the assessment.

## Stock weights

Stock weights at age have been derived from the age samples of the 3rd quarter landings since 1984 (R. Nash pers comm.). The stock mean weights for 1975-83 are time invariant and were re-examined in 1985 (Anon. 1985). They result from combining Manx and Mourne data sets. The weights at age of those stocks were considered relatively stable over time.

## Mean weights

Mean weights-at-age in the catch (1985 to 2007) are given in Table 3. Mean weights-at-age of all ages remained low. There has been a change in mean weight over the time period 1961 to the present (ICES, 2003 ACFM:17). Mean weights-at-age increased between the early 1960s and the late 1970s whereupon there has been a steady decline to the early 1990s, where they remained low. In the assessment, mean weights-at-age for the period 1972 to 1984 are taken as unchanging. In extending the data series back from 1971 to 1961, mean weights-at-age in the catch were taken from samples recorded by the Port Erin Marine Laboratory (ICES, 2003 ACFM:17).

There was some uncertainty in the mean weights-at-age for 2003 presented to the WG, and consequently the WG replaced these with the average mean stock weights-at-age for the preceding five years (1998 to 2002).

## Mean Lengths

Mean lengths-at-age are calculated using the catch data and are given for the years 1985 to 2006 in Table 4. In general, mean lengths have been relatively stable over the last few years and this trend has continued in 2006.

## Catch at length

Catch at length are listed for the years 1990-2004 (Table 5)

## B.3. Surveys

The following surveys have provided data for the VIIa(N) assessment:

| SURVEY ACRONYM | TYPE | Abundance data | Area and Month | Period |
| :---: | :---: | :---: | :---: | :---: |
| AC(VIIaN) | Acoustic survey | Numbers at age (1-ring and older); SSB | VIIa(N) from $53^{\circ} 20^{\prime} \mathrm{N}$ $55^{\circ} \mathrm{N}$; September | 1994 - present |
| NINEL | Larva survey | Production of larvae at 6mm TL | VIIa(N) from $53^{\circ} 50^{\prime} \mathrm{N}$ $54^{\circ} 50^{\prime} \mathrm{N}$; November | 1993 - present |
| DBL | Larva survey | Production of larvae at 6 mm TL | East coast of Isle of Man; October | $\begin{aligned} & 1989-1999(1996 \\ & \text { missing }) \end{aligned}$ |
| GFS-oct | Groundfis h survey | Mean nos. caught per 3 n.miles (1\&2 ringers), by region | ```VIIa(N) from 530}20'N - 54*}50`\mp@code{N (stratified); October``` | 1993-1999 |
| GFS-mar | Groundfis h survey | Mean nos. caught per 3 n.miles ( $1 \& 2$ ringers), by region | $\begin{aligned} & \text { VIIa(N) from } 53^{\circ} 20^{\prime} \mathrm{N}- \\ & 54^{\circ} 50^{\prime} \mathrm{N} \text { (stratified); } \\ & \text { March } \end{aligned}$ | 1993-1999 |

Data from a number of earlier surveys have been documented in the ICES WG reports. These include:

NW Irish Sea young herring surveys (Irish otter trawl survey using commercial trawler; 1980-1988)

Douglas Bank (East Isle of Man) larva surveys (ring net surveys; 1974-1988) (Port Erin Marine Lab)
Douglas Bank spawning aggregation acoustic surveys (1989, 1990, 1994, 1995) (Port Erin Marine Lab)

Western Irish Sea acoustic survey ( July 1991, 1992) (UK(NI))
Eastern Irish Sea acoustic survey (December 1996)
Surveys used in recent assessments are described below.
AC(VIIaN) acoustic survey
This survey uses a stratified design with systematic transects, during the first two weeks of September. Vessel currently used is the R.V. Corystes (UK(NI)) replacing the R.V. Lough Foyle (UK(NI)). Starting positions are randomized each year (see recent HAWG reports for transect design and survey results). The survey is most intense around the Isle of Man ( 2 to 4 n.mile transect spacing) where highest densities of adult herring are expected based on previous surveys and fishery data. Transect spacing of 6 to 10 n.miles are used elsewhere. A sphere-calibrated EK- 50038 kHz sounder is employed, and data are archived and analysed using Echoview (SonarData, Tasmania). Targets are identified by midwater trawling. Acoustic records are manually partitioned to species by scrutinising the echograms and using trawl compositions where appropriate. ICES-recommended target strengths are used for herring, sprat, mackerel, horse mackerel and gadoids. The survey design and implementation follows, where possible, the guidelines for ICES herring acoustic surveys in the North Sea and West of Scotland. The survey data are analysed in 15-minute elementary distance sampling units (approx. 2.5 n.miles). An estimate of density by age class, and spawning stock biomass, is obtained for each EDSU and a distance-weighted average calculated for each stratum. These are raised by stratum area to give population numbers and SSB by stratum.

## NINEL larva survey

The DARD herring larva survey has been carried out in November each year since 1993. Sampling is carried out on a systematic grid of stations covering the spawning grounds and surrounding regions in the NE and NW Irish Sea (Figure 1). Larvae are sampled using a Gulf-VII high-speed plankton sampler with $280 \mu \mathrm{~m}$ net. Doubleoblique tows are made to within 2 m of the seabed at each station. Internal and external flow rates, and temperature and salinity profiles, were recorded during each tow. Lengths of all herring larva captured are recorded.

Mean catch-rates (nos. $\mathrm{m}^{-2}$ ) are calculated over stations to give separate indices of abundance for the NE and NW Irish Sea. Larval production rates (standardised to a larva of 6 mm ), and birth-date distributions, are computed based on the mean density of larvae by length class. A growth rate of 0.35 mm day ${ }^{-1}$ and instantaneous mortality of 0.14 day $^{-1}$ are assumed based on estimates made in 1993-1997. More recent studies have indicated a mortality rate of 0.09 , and this value is also applied to examine the effect on trends in estimates of larval production

## DBL larva survey

Herring larvae were sampled on the east side of the Isle of Man in September or October each year. Double oblique tows with a 60 cm Gulf VII/PRO-NET high-speed plankton sampler with a 40 cm aperture nose cone were undertaken on a 5 Nm square grid. The tow profile was followed with a FURUNO net sonde attached to the top of
the equipment. The volume of water filtered was calculated from the nose cone mouth flow meter. The samples were preserved in $4 \%$ seawater buffered formalin and stored in 70\% alcohol.

All herring larvae were sorted from the samples. The numbers of larvae per $\mathrm{m}^{3}$ were calculated from the volume of water filtered and the number of larvae per tow. Up to 100 larvae from each tow were measured with an ocular graticule in a stereo microscope. Each sample was assigned to a sampling square and the total number of larvae per 0.5 mm size class calculated from the average depth of the square and the surface area.

The total production and time of larvae hatch was calculated using an instantaneous mortality coefficient $(\mathrm{k})$ of 0.14 and a growth rate of $0.35 \mathrm{~mm} \mathrm{~d}^{-1}$ in the formula:

$$
N_{t}=N_{o} e^{-(k t)}
$$

Production was calculated as the sum of all size classes/hatching dates. Spawning dates were taken as 10 days prior to the hatching date (Bowers 1952).

The Douglas Bank Larva survey has not been updated since 1999. Examination of the sum of squares surface from SPALY in 2005 indicated that the Douglas Bank larvae index (DBL) was having no influence in the assessment estimates for the current year. Therefore, the WG agreed on removing DBL from the analysis (ICES, 2005). The DBL time series is listed in Table 6

## GFS-oct and -mar groundfish surveys

The DARD groundfish survey of ICES Division VIIaN are carried out in March and October at standard stations between $53^{\circ} 20^{\prime} \mathrm{N}$ and $54^{\circ} 45^{\prime} \mathrm{N}$ (Figure 2). Data from additional stations fished in the St George's Channel since October 2001 have not been used in calculating herring indices of abundance. As in previous surveys, the area was divided into strata according to depth contour and sediment type, with fixed station positions (note that the strata in Fig. 2 differ from those in the September acoustic survey shown in Fig. 1). The sampling gear was a Rockhopper otter trawl fitted with non-rotating rubber discs of approximately 15 cm diameter on the footrope. The trawl fishes with an average headline height of 3.0 m and door spread of 30 - 40 m depending on depth and tide. A 20 mm stretched-mesh codend liner was fitted. During March, trawling was carried out at an average speed of 3 knots across the ground, over a standard distance of 3 nautical miles at standard stations and 1 nautical mile in the St. George's Channel. Since 2002, all survey stations in the October survey have been of 1-mile distance. Comparative trawling exercises during the October surveys and during an independent exercise in February 2003 indicate roughly similar catch-rates per mile between 1-mile and 3-mile tows. It is planned to continue with some comparative trawling experiments during future surveys to improve the statistical power of significance tests between the 1-mile and 3-mile tows.

As the surveys are targeted at gadoids, ages were not recorded for herring. The length frequencies in each survey were sliced into length ranges corresponding to 0 ring and 1-ring herring according to the appearance of modes in the overall weighted mean length frequency for each survey. Some imprecision will have resulted because of the overlap in length-at-age distributions of 1-ring and 2-ring herring. The error is considered to be comparatively small for most of the surveys where clear modes are apparent. There was no clear division between 1-ring and 2-ring herring in the March 2003 groundfish survey, and the estimate for 1-ringers may include a significant component of small 2-ringers. The arithmetic mean catch-rate and approximate vari-
ance of the mean was computed for each age-class in each survey stratum, and averaged over strata using the areas of the strata as weighting factors.

Groundfish surveys were used by the 1996 to 1999 HAWG to obtain indices for 0-and 1-ring herring in the Irish Sea. These indices have performed poorly in the assessment and have not been used since 1999. The time-series is listed in Table 7.

## B.4. Commercial CPUE

Commercial CPUE's are not used for this stock.
B.5. Other relevant data
C. Historical Stock Development

Model used: ICA
Software used: ICA (Patterson 1998)
Model Options chosen:
Separable constraint over last 6 years (weighting $=1.0$ for each year)
Reference age $=4$
Constant selection pattern model
Selectivity on oldest age $=1.0$
First age for calculation of mean $\mathrm{F}=2$
Last age for calculation of mean $\mathrm{F}=6$
Weighting on 1-rings $=0.1$; all other age classes $=1.0$
Weighting for all years $=1.0$
All indices treated as linear
No S/R relationship fitted
Lowest and highest feasible $\mathrm{F}=0.05$ and 2.0
All survey weights fitted by hand i.e., 1.0 with the 1 ringers in the acoustic survey weighted to 0.1.

Correlated errors assumed i.e., $=1.0$
No shrinkage applied

Input data types and characteristics:

| TYPE | Name | Year range | Age range | Variable from year to year Yes/No |
| :---: | :---: | :---: | :---: | :---: |
| Caton | Catch in tonnes | 1961-last data year | NA | Yes |
| Canum | Catch at age in numbers | 1961-last data year | 1-8+ | Yes |
| Weca | Weight at age in the commercial catch | $\begin{aligned} & \hline \text { 1961-1971 } \\ & \text { 1972-1983 } \\ & \text { 1984-last data } \\ & \text { year } \\ & \hline \end{aligned}$ | $\begin{aligned} & 1-8+ \\ & 1-8+ \\ & 1-8+ \end{aligned}$ | $\begin{aligned} & \text { Yes } \\ & \text { No } \\ & \text { Yes } \end{aligned}$ |
| West | Weight at age of the spawning stock at spawning time. | $\begin{array}{\|l\|} \hline \text { 1961-1971 } \\ \text { 1972-1983 } \\ \text { 1984-last data } \\ \text { year } \\ \hline \end{array}$ | $\begin{aligned} & 1-8+ \\ & 1-8+ \\ & 1-8+ \end{aligned}$ | Yes <br> No <br> Yes |
| Mprop | Proportion of natural mortality before spawning | 1961-last data year | NA | No |
| Fprop | Proportion of fishing mortality before spawning | 11961-last data year | NA | No |
| Matprop | Proportion mature at age | 1961-last data year | 1-8+ | Yes |
| Natmor | Natural mortality | 1961-last data year | 1-8+ | No |

Tuning data:

| TYPE | NAME | Year range | AGE RANGE |
| :--- | :--- | :--- | :--- |
| Tuning fleet 1 | NINEL | $1993-2003$ | SSB |
| Tuning fleet 2 | DBL | $1989-1999$ | SSB |
| Tuning fleet 3 | GFS-octtot | $1993-2005$ | $1 \& 2$ |
| Tuning fleet 4 | GFS-martot | $1992-2003$ | 1 |
| Tuning fleet 5 | ACAGE | $1994-2003$ | $1-8+$ |
| Tuning fleet 6 | AC_VIIa(N) | $1994-2003$ | SSB |
| Tuning fleet 7 | AC_1+ | $1994-2003$ | SSB/Total biomass |

## Two-stage biomass model

In 2005 a Two-Stage Biomass model for the assessment of Irish Sea VIIa(N) herring given additional variance in the recruitment index was presented by Roel and De Oliveira (ICES 2005 WD10).

The model addresses the problem of the high uncertainty in the assessment of Irish Sea herring, which to some extent may be related to the presence of juvenile Celtic Sea herring in both the fishery and the survey area. In the absence of a Celtic Sea herring recruitment index, the biomass model presented addressed the problem by limiting recruitment variability in Irish Sea herring on the basis of information available for other herring stocks. The total variability in the recruitment data was divided into two components: the one related to Irish Sea herring recruitment variability and the
rest which was likely to represent variability related to the presence of Celtic Sea juveniles.

The model is fitted to biomass indices of 1-ringer fish and to aggregated biomass indices for the 2-rings+ from Northern Ireland acoustic surveys. The survey age composition data and the weights-at-age from the catch are used to calculate the proportion of 1-ring fish in the survey. The proportion is then applied to the total acoustic biomass to compute the 1 -ring biomass index while the 2 -ring+ index is obtained by subtraction. The catch in weight was split in a similar manner but based on commercial catch samples.

## The model

The dynamics take into account only two stages in the population: the recruits, 1ringer fish, and the fully recruited that comprise 2-ringer and older fish. The biomass dynamics is represented by the following:

$$
\begin{equation*}
B_{y+1}=B_{1, y+1}+\left[\left(B_{2+, y}+B_{1, y}\right) e^{-3 g / 4}-C_{y}\right] e^{-g / 4} \tag{1}
\end{equation*}
$$

where
$B_{1, y} \quad$ is the biomass of recruitment (tons) at the start of year $y$;
$B_{2+y} \quad$ is the biomass of $2+$ aged fish (tons) at the start of year $y$;
$C_{y} \quad$ is the biomass of fish caught (tons) during year $y$, assumed to be taken in a pulse fishery $3 / 4$ of the way into year $y$; and
$g \quad$ is a composite parameter, treated as an annual rate, which accounts for natural mortality and growth.

Maximum likelihood estimation is used, assuming survey indices are log-normally distributed about their expected values. Standard errors of the log-distributions are approximated by the sampling CVs of the untransformed distributions.

The estimable parameters are $g, B_{2+, 1994}, B_{1,1994}, \ldots, B_{1,2004}, \lambda^{2}$ and $q$
where $q$ corresponds to the catchability associated with the survey indices $I_{1, y}$ and $I_{2+y}$ and $\lambda^{2}$ is the additional variance.

The data were explored for values of recruitment variability $\left(\sigma_{R}\right)=0.4$ and 0.8 . The value 0.4 corresponds to the variability in recruitment age 1 as estimated by ICA for the period used in this analysis, but excluding the most recent estimate (1994-2006). The two parameters, $g$ and $q$, may be confounded in the model indicating that fixing $g$ was appropriate. This parameter was fixed to 0.2 following a similar approach as in Roel and De Oliveira (ICES 2005 WD10).

## D. Short-Term Projection

NOT USED IN 2004
Model used: Age structured
Software used: MFDP ver 1a
Initial stock size: Taken from the last year of the assessment. 1-ring recruits taken from a geometric mean for the years 1983 to two years prior to the current year. Where 1-ringers are absurdly estimated in the assessment 2-ringers are estimated as a geometric mean of the previous 10 year period.

Maturity: Mean of the previous three years of the maturity ogive used in the assessment.

F and M before spawning: Set to 0.9 and 0.75 respectively for all years.
Weight at age in the stock: Mean of the previous three years in the assessment.
Weight at age in the catch: Mean of the previous three years in the assessment.
Exploitation pattern: Mean of the previous three years, scaled by the Fbar (2-6) to the level of the last year.

Intermediate year assumptions: TAC constraint.
Stock recruitment model used: None used
Procedures used for splitting projected catches: Not relevant

## E. Medium-Term Projections

## F. Long-Term Projections

Not done

## G. Biological Reference Points

Until there is confidence in the assessment the Working Group decided not to revisit the estimation of $\mathbf{B}_{\mathrm{pa}}(9,500 \mathrm{t})$ and $\mathbf{B}_{\lim }(6,000 \mathrm{t})$. There were no new points to add to the discussions and deliberations presented in 2000 (ICES 2000/ACFM:10).

## H. Other Issues

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## Table 1. Biological sampling of Irish Sea (VIIa(N)) landings. Country denotes sampling nation.

|  | Coverage | $\%$ of landings sampled | No of samples | Total landings | landings by Q? | IRELAND |  |  |  | NORTHERN | N IRELAND |  |  | ISLE OF MA |  |  |  | OTHERr UK | UK OFFS | HORE |  | TOTAL |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year |  |  |  |  |  | Landings | Samples | Lengths | Ages | Landings | Samples | Lengths | Ages | Landings | Samples | Lengths | Ages | Landings | Samples | Lengths | Ages | Landings | Samples | Lengths | Ages |
| 1988 | (4) |  |  |  |  | **2579 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 0 |
| 1989 | (3) temp spread good |  | 88 | 4962 | NO | 1430 | 21 | 1843 | 555 |  | 45 | 11464 | 2249 |  | 21 | 5173 | 1057 |  | 1 | 96 | 0 | 4962 | 88 | 18576 | 3861 |
| 1990 | $\mathrm{p}(1,2)$ | 68\% | 100 | 6312 | YES | 1699 | 44 | 5176 | 1022 | 2322 | 38 | 9310 | 1900 | 542 | 18 | 5276 | 897 | 179/1570 | 0 | 0 | 0 | 6312 | 100 | 19762 | 3819 |
| 1991 | g | - $90 \%$ | 138 | 4398 | YES | 80 | 5 | 1255 | 247 | 3298 | 105 | 16724 | 2484 | 629 | 28 | 8280 | 1392 | 0/391 | 0 | 0 | 0 | 4398 | 138 | 26259 | 4123 |
| 1992 | g | 98\% | 32 | 5270 | YES | 406 | 3 | 593 | 99 | 4120 | 16 | 1588 | 770 | 741 | 13 | 3488 | 680 | 3 | 0 | 0 | 0 | 5270 | 32 | 5669 | 1549 |
| 1993 | p (1) | 65\% | 48 | 4408 | YES | 0 | 5 | 1378 | 245 | 3632 | 34 | 3744 | 832 | 776 | 9 | 1560 | 448 | 0 | 0 | 0 | 0 | 4408 | 48 | 6682 | 1525 |
| 1994 | v.g | 95\% | 59 | 4828 | YES | 0 | 21 | 569 | 100 | 3956 | 43 | 3691 | 1175 | 716 | 14 | 3724 | 614 | 156 | 0 | 0 | 0 | 4828 | 59 | 7984 | 1889 |
| 1995 | $\mathrm{g}(1)$ | 87\% | 85 | 5076 | YES | 0 | 21 | 569 | 100 | 3860 | 75 | 8282 | 2545 | 615 | 8 | 2182 | 400 | 601 | 0 | 0 | 0 | 5076 | 85 | 11033 | 3045 |
| 1996 | $\mathrm{g}(1,5)$ | 70\% | 51 | 5301 | YES | 100 | 1 | 537 | 55 | 4335 | 45 | 4813 | 1050 | 537 | 5 | 997 | 228 | 329 | 0 | 0 | 0 | 5301 | 51 | 6347 | 1333 |
| 1997 | $\mathrm{g}(1,2)$ | 91\% | 34 | 6649 | YES | 0 | 2 | 473 | 50 | 5679 | 25 | 2900 | 1199 | 765 | 7 | 2246 | 340 | 205 | 0 | 234 | 76 | 6649 | 34 | 5853 | 1665 |
| 1998 | g (2) | 84\% | 31 | 4904 | YES | 0 | 2 | 150 | 50 | 4131 | 29 | 2979 | 1450 | 0 | 0 | 0 | 0 | 7732 | 0 | 0 | 0 | 4904 | 31 | 3129 | 1500 |
| 1999 | g (2) | 72\% | 32 | 4127 | YES | 0 | 4 | 0 | 200 | 2967 | 28 | 2518 | 1400 | 0 | 0 | 0 | 0 | 11602 | 0 | 0 | 0 | 4127 | 32 | 2518 | 1600 |
| 2000 | v.g | 97\% | 28 | 2002 | YES | 0 | 5 | 932 | 0 | 2002 | 23 | 1915 | 1150 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2002 | 28 | 2847 | 1150 |
| 2001 | p (2) | 70\% | 31 | 5461 | YES | 862 | 8 | 1031 | 222 | 3786 | 23 | 2915 | 1149 | 86 | 0 | 0 | 0 | 7272 | 0 | 0 | 0 | 5461 | 31 | 3946 | 1371 |
| 2002 | p (1) | 62\% | 9 | 2392 | YES | 286 | 0 | 0 | 0 | 2051 | 9 | 949 | 450 | 4 | 0 | 0 | 0 | 51 | 0 | 0 | 0 | 2392 | 9 | 949 | 450 |
| 2003 |  |  | 9 | 2399 | YES | 0 |  |  |  | 2399 | 9 | 1132 | 445 |  |  |  |  |  |  |  |  |  |  |  |  |
| 2004 |  |  | 9 | 2531 | YES | 749 | 2 | 190 | 133 | 1782 | 7 | 991 | 350 |  |  |  |  |  |  |  |  |  |  |  |  |
| 2005 |  |  | 26 | 4387 | YES | 1153 | 5 | 1312 | 372 | 3234 | 21 | 4135 | 1018 |  |  |  |  |  |  |  |  |  |  |  |  |
| 2006 |  |  | 22 | 4402 | YES | 581 | 8 | 2248 | 549 | 3821 | 14 | 1982 | 686 |  |  |  |  |  |  |  |  |  |  |  |  |
| 2007 |  |  | 29 | 4629 | YES | 0 |  |  |  | 4629 | 29 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2008 |  |  | 19 | 4895 | YES | 0 |  |  |  | 4895 | 19 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

 related to this level of detail:
VERY GOOD (v.g) : all landings which individually are $>10 \%$ of the total were sampled, all $Q$ for which there were landings were sampled
GOOD (g) : landings that constitute the majority of the catch (adding to approx $70 \%$ or more of total) were sampled
POOR (p)
(1): unsampled quarters
: some of the large landings not sampled
(2): large landings with few samples or unsampled. High level of sampling corresponds to 1 sample per 100t landed (WG rep 1997)
(3): Comment from WG rep. From 1990 going back, Report landings and sampling levels are shown aggregated for the whole year. UK landings lumped in one figure.
 labs.
(5): NO samples for NI landings in 4th $Q$, there is a suspicion that the figures correspond to 'paper landings'.
${ }^{1}$ Samples applied to NI landings: ${ }^{2}$ Large unsampled landings.

Table 2: Data and method used to estimate landings from Division VIIa(N) herring.




Figure 1. Sampling stations for larvae in the North Irish Sea (NINEL). Sampling is undertaken in November each year.


Key to strata: 1. Irish Coast (N), $<100 \mathrm{~m}$, Mixed sediments
2. Irish Coast, $<50 \mathrm{~m}$, sand and finer sediments
3. Irish Coast, 50-100m, Muddy sediments
4. W and SW Isle of Man, 50-100m, mud and muddy sand
5. N Isle of Man, $<50 \mathrm{~m}$, gravel sediments
6. Eastern Irish Sea, $<50 \mathrm{~m}$, sand and finer sediments
7. S. Isle of Man, $<100 \mathrm{~m}$, gravel sediments
8. Deep western channel and North Channel $>100 \mathrm{~m}$
9. St George's Channel west; sandy/mixed sediments; $<100 \mathrm{~m}$
10. St George's Channel east; sandy/mixed sediments; $<100 \mathrm{~m}$

Figure 2. Standard station positions for DARD groundfish survey of the Irish Sea in March and October. Boundaries of survey strata are shown. Indices for the "Western Irish Sea" use data from strata 2-4. Indices for the "Eastern Irish Sea" use data from stratum 6 only (few juvenile herring are found in stratum 7). (Note different stratification to Fig. 1.). New stations fished in the St Georges Channel (strata 9 and 10) since October 2001 are not included in the survey indices. Stratum 5 ( 1 station only in recent years) is also excluded from the index. There are no stations in stratum 8 due to difficult trawling conditions for the gear used in the survey. Station 121 in stratum 7 has been fished only once and is excluded from the index.

Table 3. Irish Sea Herring Division VIIa(N). Mean weights-at-age in the catch.

| Year | Weights-at-age (g) Age (rings) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1985 | 87 | 125 | 157 | 186 | 202 | 209 | 222 | 258 |
| 1986 | 68 | 143 | 167 | 188 | 215 | 229 | 239 | 254 |
| 1987 | 58 | 130 | 160 | 175 | 194 | 210 | 218 | 229 |
| 1988 | 70 | 124 | 160 | 170 | 180 | 198 | 212 | 232 |
| 1989 | 81 | 128 | 155 | 174 | 184 | 195 | 205 | 218 |
| 1990 | 77 | 135 | 163 | 175 | 188 | 196 | 207 | 217 |
| 1991 | 70 | 121 | 153 | 167 | 180 | 189 | 195 | 214 |
| 1992 | 61 | 111 | 136 | 151 | 159 | 171 | 179 | 191 |
| 1993 | 88 | 126 | 157 | 171 | 183 | 191 | 198 | 214 |
| 1994 | 73 | 126 | 154 | 174 | 181 | 190 | 203 | 214 |
| 1995 | 72 | 120 | 147 | 168 | 180 | 185 | 197 | 212 |
| 1996 | 67 | 116 | 148 | 162 | 177 | 199 | 200 | 214 |
| 1997 | 64 | 118 | 146 | 165 | 176 | 188 | 204 | 216 |
| 1998 | 80 | 123 | 148 | 163 | 181 | 177 | 188 | 222 |
| 1999 | 69 | 120 | 145 | 167 | 176 | 188 | 190 | 210 |
| 2000 | 64 | 120 | 148 | 168 | 188 | 204 | 200 | 213 |
| 2001 | 67 | 106 | 139 | 156 | 168 | 185 | 198 | 205 |
| 2002 | 85 | 113 | 144 | 167 | 180 | 184 | 191 | 217 |
| 2003* | 81 | 116 | 136 | 160 | 167 | 172 | 186 | 199 |
| 2004 | 73 | 107 | 130 | 157 | 165 | 187 | 200 | 205 |
| 2005 | 67 | 103 | 136 | 156 | 166 | 180 | 191 | 209 |
| 2006 | 64 | 105 | 131 | 149 | 164 | 177 | 184 | 211 |
| 2007 | 67 | 112 | 135 | 158 | 173 | 183 | 199 | 227 |
| 2008 | 71 | 110 | 135 | 153 | 156 | 182 | 196 | 206 |

[^9]Table 4. Irish Sea Herring Division VIIa(N). Mean length-at-age in the catch.

| Year | Lengths-at-age (cm) Age (rings) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| 1985 | 22.1 | 24.3 | 26.1 | 27.6 | 28.3 | 28.6 | 29.5 | 30.1 |
| 1986 | 19.7 | 24.3 | 25.8 | 26.9 | 28.0 | 28.8 | 28.8 | 29.8 |
| 1987 | 20.0 | 24.1 | 26.3 | 27.3 | 28.0 | 29.2 | 29.4 | 30.1 |
| 1988 | 20.2 | 23.5 | 25.7 | 26.3 | 27.2 | 27.7 | 28.7 | 29.6 |
| 1989 | 20.9 | 23.8 | 25.8 | 26.8 | 27.8 | 28.2 | 28.0 | 29.5 |
| 1990 | 20.1 | 24.2 | 25.6 | 26.2 | 27.7 | 28.3 | 28.3 | 29.0 |
| 1991 | 20.5 | 23.8 | 25.4 | 26.1 | 26.8 | 27.3 | 27.7 | 28.7 |
| 1992 | 19.0 | 23.7 | 25.3 | 26.2 | 26.7 | 27.2 | 27.9 | 29.4 |
| 1993 | 21.6 | 24.1 | 25.9 | 26.7 | 27.2 | 27.6 | 28.0 | 28.7 |
| 1994 | 20.1 | 23.9 | 25.5 | 26.5 | 27.0 | 27.4 | 27.9 | 28.4 |
| 1995 | 20.4 | 23.6 | 25.2 | 26.3 | 26.8 | 27.0 | 27.6 | 28.3 |
| 1996 | 19.8 | 23.5 | 25.3 | 26.0 | 26.6 | 27.6 | 27.6 | 28.2 |
| 1997 | 19.6 | 23.6 | 25.1 | 26.0 | 26.5 | 27.1 | 27.7 | 28.2 |
| 1998 | 20.8 | 23.8 | 25.2 | 26.1 | 27.0 | 26.8 | 27.2 | 28.7 |
| 1999 | 19.8 | 23.6 | 25.0 | 26.1 | 26.5 | 27.1 | 27.2 | 28.0 |
| 2000 | 19.7 | 23.8 | 25.3 | 26.3 | 27.1 | 27.7 | 27.7 | 28.1 |
| 2001 | 20.0 | 22.9 | 24.8 | 25.7 | 26.2 | 26.9 | 27.5 | 27.8 |
| 2002 | 21.1 | 23.1 | 24.8 | 26.0 | 26.6 | 26.7 | 27.0 | 28.1 |
| 2003 | 21.1 | 23.7 | 25.0 | 26.5 | 26.9 | 27.1 | 27.8 | 28.5 |
| 2004 | 20.7 | 23.1 | 24.6 | 25.8 | 26.1 | 27.1 | 27.6 | 28.3 |
| 2005 | 20.0 | 22.6 | 24.5 | 25.5 | 26.0 | 26.6 | 27.1 | 27.8 |
| 2006 | 19.5 | 22.7 | 24.3 | 25.3 | 26.0 | 26.6 | 26.9 | 28.0 |
| 2007 | 20.1 | 23.0 | 24.1 | 25.1 | 25.8 | 26.2 | 26.7 | 27.8 |
| 2008 | 20.0 | 22.7 | 24.1 | 25.0 | 25.2 | 26.3 | 26.9 | 27.4 |

Table 5. Irish Sea Herring Division VIIa (N). Catch-at-length for 1990-2004. Numbers of fish in thousands.

| Length | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 14.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 15 |  |  | 95 |  |  |  |  |  |  |  |  |  |  |  |  |
| 15.5 |  |  | 169 |  |  |  |  |  |  | 10 |  |  |  |  |  |
| 16 | 6 |  | 343 |  |  | 21 | 21 | 17 |  | 19 | 12 | 9 |  |  |  |
| 16.5 | 6 | 2 | 275 |  |  | 55 | 51 | 94 |  | 53 | 49 | 27 |  |  | 13 |
| 17 | 50 | 1 | 779 |  | 84 | 139 | 127 | 281 | 26 | 97 | 67 | 53 |  |  | 25 |
| 17.5 | 7 | 4 | 1106 |  | 59 | 148 | 200 | 525 | 30 | 82 | 97 | 105 |  |  | 84 |
| 18 | 224 | 31 | 1263 |  | 69 | 300 | 173 | 1022 | 123 | 145 | 115 | 229 |  |  | 102 |
| 18.5 | 165 | 56 | 1662 |  | 89 | 280 | 415 | 1066 | 206 | 135 | 134 | 240 | 36 |  | 114 |
| 19 | 656 | 168 | 1767 | 39 | 226 | 310 | 554 | 1720 | 317 | 234 | 164 | 385 | 18 |  | 203 |
| 19.5 | 318 | 174 | 1189 | 75 | 241 | 305 | 652 | 1263 | 277 | 82 | 97 | 439 | 0 | 29 | 269 |
| 20 | 791 | 454 | 1268 | 75 | 253 | 326 | 749 | 1366 | 427 | 218 | 109 | 523 | 0 | 73 | 368 |
| 20.5 | 472 | 341 | 705 | 57 | 270 | 404 | 867 | 1029 | 297 | 242 | 85 | 608 | 18 | 215 | 444 |
| 21 | 735 | 469 | 705 | 130 | 400 | 468 | 886 | 1510 | 522 | 449 | 115 | 1086 | 307 | 272 | 862 |
| 21.5 | 447 | 296 | 597 | 263 | 308 | 782 | 1258 | 1192 | 549 | 362 | 138 | 1201 | 433 | 290 | 1007 |
| 22 | 935 | 438 | 664 | 610 | 700 | 1509 | 1530 | 2607 | 1354 | 1261 | 289 | 1748 | 1750 | 463 | 1495 |
| 22.5 | 581 | 782 | 927 | 1224 | 785 | 2541 | 2190 | 2482 | 1099 | 2305 | 418 | 1763 | 1949 | 600 | 2140 |
| 23 | 2400 | 1790 | 1653 | 2016 | 1035 | 4198 | 2362 | 3508 | 2493 | 4784 | 607 | 2670 | 2490 | 1158 | 2089 |
| 23.5 | 1908 | 1974 | 1156 | 2368 | 1473 | 4547 | 2917 | 3902 | 2041 | 4183 | 951 | 2254 | 1552 | 1380 | 2214 |
| 24 | 3474 | 2842 | 1575 | 2895 | 2126 | 4416 | 3649 | 4714 | 3695 | 4165 | 1436 | 3489 | 1029 | 1273 | 2054 |
| 24.5 | 2818 | 2311 | 2412 | 2616 | 2564 | 3391 | 4077 | 4138 | 2769 | 3397 | 1783 | 4098 | 758 | 1249 | 2269 |
| 25 | 4803 | 2734 | 2792 | 2207 | 3315 | 3100 | 4015 | 5031 | 2625 | 2620 | 2144 | 5566 | 776 | 1163 | 1749 |
| 25.5 | 3688 | 2596 | 3268 | 2198 | 3382 | 2358 | 3668 | 3971 | 2797 | 1817 | 1791 | 4785 | 1335 | 1211 | 1206 |
| 26 | 4845 | 3278 | 3865 | 2216 | 3480 | 2334 | 2480 | 3871 | 3115 | 1694 | 1349 | 3814 | 1570 | 1140 | 823 |
| 26.5 | 3015 | 2862 | 3908 | 2176 | 2617 | 1807 | 2177 | 2455 | 2641 | 1547 | 840 | 2243 | 1552 | 1573 | 587 |
| 27 | 3014 | 2412 | 3389 | 2299 | 2391 | 1622 | 1949 | 1711 | 2992 | 1475 | 616 | 1489 | 776 | 1607 | 510 |
| 27.5 | 1134 | 1449 | 2203 | 2047 | 1777 | 990 | 1267 | 1131 | 1747 | 867 | 479 | 644 | 433 | 1189 | 383 |
| 28 | 993 | 922 | 1440 | 1538 | 1294 | 834 | 906 | 638 | 1235 | 276 | 212 | 496 | 162 | 726 | 198 |
| 28.5 | 582 | 423 | 569 | 944 | 900 | 123 | 564 | 440 | 170 | 169 | 58 | 179 | 108 | 569 | 51 |
| 29 | 302 | 293 | 278 | 473 | 417 | 248 | 210 | 280 | 111 | 61 | 42 | 10 | 36 | 163 |  |
| 29.5 | 144 | 129 | 96 | 160 | 165 | 56 | 79 | 59 | 92 |  | 12 | 0 | 36 | 129 |  |
| 30 | 146 | 82 | 70 | 83 | 9 | 40 | 32 | 8 | 84 |  | 6 | 9 |  | 43 |  |
| 30.5 | 57 | 36 | 36 | 15 | 27 | 5 | 0 | 5 | 3 |  |  |  |  | 43 |  |
| 31 | 54 | 12 | 2 | 4 |  | 1 | 2 |  |  |  |  |  |  | 43 |  |
| 31.5 | 31 | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 32 | 29 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 32.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 33 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 33.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 34 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 5 (continued). Irish Sea Herring Division VIIa (N). Catch-at-length for 1990-2004. Numbers of fish in thousands.

| Length | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: |
| 14 |  |  |  |  |
| 14.5 |  |  |  |  |
| 15 |  |  |  |  |
| 15.5 |  |  | 16 |  |
| 16 |  | 2 |  |  |
| 16.5 | 1 | 44 | 33 | 1 |
| 17 | 39 | 140 | 69 | 3 |
| 17.5 | 117 | 211 | 286 | 11 |
| 18 | 291 | 586 | 852 | 34 |
| 18.5 | 521 | 726 | 2088 | 64 |
| 19 | 758 | 895 | 2979 | 85 |
| 19.5 | 933 | 1246 | 3527 | 108 |
| 20 | 943 | 984 | 3516 | 100 |
| 20.5 | 923 | 1443 | 2852 | 133 |
| 21 | 1256 | 1521 | 3451 | 192 |
| 21.5 | 1380 | 1621 | 2929 | 217 |
| 22 | 1361 | 2748 | 3821 | 271 |
| 22.5 | 1448 | 3629 | 3503 | 229 |
| 23 | 1035 | 4358 | 4196 | 322 |
| 23.5 | 1256 | 2920 | 3697 | 264 |
| 24 | 1276 | 3679 | 3178 | 259 |
| 24.5 | 1083 | 2431 | 2136 | 204 |
| 25 | 1086 | 3438 | 1503 | 148 |
| 25.5 | 584 | 2198 | 952 | 114 |
| 26 | 438 | 1714 | 643 | 78 |
| 26.5 | 203 | 605 | 330 | 42 |
| 27 | 165 | 445 | 147 | 23 |
| 27.5 | 60 | 155 | 72 | 10 |
| 28 | 45 | 104 | 33 | 12 |
| 28.5 | 18 | 9 | 26 | 1 |
| 29 | 12 | 46 |  |  |
| 29.5 |  |  | 7 |  |
| 30 |  |  |  |  |
| 30.5 |  |  |  |  |
| 31 |  |  |  |  |
| 31.5 |  |  |  |  |
| 32 |  |  |  |  |
| 32.5 |  |  |  |  |
| 33 |  |  |  |  |
| 33.5 |  |  |  |  |
| 34 |  |  |  |  |

Table 6. Irish Sea herring Division VIIa(N). Northern Ireland groundfish survey indices for herring (Nos. per 3 miles).
(a) 0-ring herring: October survey

|  | Western Irish Sea |  |  | EASTERN IRISH SEA |  |  | Total IRISH SEA |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey | Mean | N.obs | SE | Mean | N.obs. | SE | Mean | N. obs | SE |
| 1991 | 54 | 34 | 22 |  |  |  |  |  |  |
| 1992 | 210 | 31 | 99 | 240 | 8 | 149 | 177 | 46 | 68 |
| 1993 | 633 | 26 | 331 | 498 | 10 | 270 | 412 | 44 | 155 |
| 1994 | 548 | 26 | 159 | 8 | 7 | 5 | 194 | 41 | 55 |
| 1995 | 67 | 22 | 23 | 35 | 9 | 18 | 37 | 35 | 11 |
| 1996 | 90 | 26 | 58 | 131 | 9 | 79 | 117 | 42 | 50 |
| 1997 | 281 | 26 | 192 | 68 | 9 | 42 | 138 | 43 | 70 |
| 1998 | 980 | 26 | 417 | 12 | 9 | 10 | 347 | 43 | 144 |
| 1999 | 389 | 26 | 271 | 90 | 9 | 29 | 186 | 43 | 96 |
| 2000 | 202 | 24 | 144 | 367 | 9 | 190 | 212 | 38 | 89 |
| 2001 | 553 | 26 | 244 | 236 | 11 | 104 | 284 | 45 | 93 |
| 2002 | 132 | 26 | 84 | 18 | 11 | 10 | 63 | 45 | 31 |
| 2003 | 1203 | 26 | 855 | 75 | 11 | 47 | 446 | 45 | 296 |
| 2004 | 838 | 26 | 292 | 447 | 11 | 191 | 469 | 45 | 125 |
| 2005 | 1516 | 26 | 1036 | 256 | 11 | 152 | 627 | 45 | 363 |
| 2006 | 4677 | 26 | 2190 | 2140 | 11 | 829 | 2468 | 45 | 822 |

(b) 1-ring herring: March Surveys.

|  | Western IRISH SEA |  |  | EASTERN IRISH SEA |  |  |  | Total IRISH Sea |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey | Mean | N.obs | SE | Mean | N.obs. | SE | Mean | N.obs | SE |  |
| 1992 | 392 | 20 | 198 | 115 | 10 | 73 | 190 | 34 | 77 |  |
| 1993 | 1755 | 27 | 620 | 175 | 10 | 66 | 681 | 45 | 216 |  |
| 1994 | 2472 | 25 | 1852 | 106 | 9 | 51 | 923 | 39 | 641 |  |
| 1995 | 1299 | 26 | 679 | 73 | 8 | 32 | 480 | 42 | 235 |  |
| 1996 | 1055 | 22 | 638 | 285 | 9 | 164 | 487 | 39 | 230 |  |
| 1997 | 1473 | 26 | 382 | 260 | 9 | 96 | 612 | 43 | 137 |  |
| 1998 | 3953 | 26 | 1331 | 250 | 9 | 184 | 1472 | 43 | 466 |  |
| 1999 | 5845 | 26 | 1860 | 736 | 9 | 321 | 2308 | 42 | 655 |  |
| 2000 | 2303 | 26 | 853 | 546 | 10 | 217 | 1009 | 44 | 306 |  |
| 2001 | 3518 | 26 | 916 | 1265 | 11 | 531 | 1763 | 45 | 381 |  |
| $2002^{\text {a }}$ | 2255 | 25 | 845 | 185 | 11 | 84 | 852 | 44 | 294 |  |
| $2002^{\text {b }}$ | 7870 | 26 | 5667 | 185 | 11 | 84 | 2794 | 45 | 1960 |  |
| 2003 | 2103 | 26 | 876 | 896 | 11 | 604 | 1079 | 45 | 382 |  |
| 2004 | 6611 | 25 | 2726 | 491 | 11 | 163 | 2486 | 44 | 945 |  |
| 2005 | 7274 | 26 | 3097 | 1240 | 8 | 375 | 3001 | 42 | 1121 |  |
| 2006 | 4249 | 26 | 1687 | 2630 | 11 | 813 | 2496 | 45 | 662 |  |

[^10]Table 6. (Continued) Irish Sea herring Division VIIa(N). Northern Ireland groundfish survey indices for herring (Nos. per 3 miles.).
(c) 1-ring herring: October Surveys

|  | Western IRISH SEA |  |  |  | EASTERN IRISH SEA |  |  | Total IRISH SEA |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: |
| Survey | Mean | N.obs | SE | Mean | N.obs. | SE | Mean | N.obs | SE |  |
| 1991 | 102 | 34 | 34 | n/a | n/a | n/a | n/a | n/a | $\mathrm{n} / \mathrm{a}$ |  |
| 1992 | 36 | 31 | 18 | 20 | 8 | 11 | 21 | 46 | 8 |  |
| 1993 | 122 | 26 | 66 | 4 | 10 | 2 | 44 | 44 | 23 |  |
| 1994 | 490 | 26 | 137 | 17 | 6 | 10 | 176 | 40 | 47 |  |
| 1995 | 153 | 22 | 61 | 3 | 9 | 1 | 55 | 35 | 21 |  |
| 1996 | 30 | 26 | 13 | 2 | 9 | 1 | 11 | 42 | 5 |  |
| 1997 | 612 | 26 | 369 | 0.2 | 9 | 0.2 | 302 | 43 | 156 |  |
| 1998 | 39 | 26 | 15 | 13 | 9 | 10 | 53 | 43 | 35 |  |
| 1999 | 81 | 26 | 41 | 104 | 9 | 95 | 74 | 43 | 40 |  |
| 2000 | 455 | 24 | 250 | 74 | 9 | 52 | 579 | 38 | 403 |  |
| 2001 | 1412 | 26 | 641 | 5 | 11 | 3 | 513 | 45 | 223 |  |
| 2002 | 370 | 26 | 111 | 4 | 11 | 2 | 291 | 45 | 158 |  |
| 2003 | 314 | 26 | 143 | 410 | 11 | 350 | 267 | 45 | 144 |  |
| 2004 | 710 | 26 | 298 | 103 | 11 | 74 | 299 | 45 | 108 |  |
| 2005 | 3217 | 25 | 1467 | 18 | 11 | 12 | 1121 | 44 | 507 |  |
| 2006 | 1458 | 26 | 669 | 40 | 11 | 18 | 523 | 45 | 231 |  |

Table 7. Irish Sea Herring Division VIIa (N). Larval production ( $10^{11}$ ) indices for the Manx component.

| Year | Douglas Bank |  |  |
| :--- | :--- | :--- | :--- |
|  | Date | Isle of Man <br> Production | SE |
| 1989 | 26 Oct | 3.39 | 1.54 |
| 1990 | 19 Oct | 1.92 | 0.78 |
| 1991 | 15 Oct | 1.56 | 0.73 |
| 1992 | 16 Oct | 15.64 | 2.32 |
| 1993 | 19 Oct | 4.81 | 0.77 |
| 1994 | 13 Oct | 7.26 | 2.26 |
| 1995 | 15 Oct | 1.58 | 1.68 |
| 1996 | 6 Nov | 5.59 | 1.25 |
| 1997 | 25 Oct | 2.27 | 1.43 |
| 1998 |  | 3.87 | 0.88 |

## Annex 9 Stock Annex Sprat in the North Sea

Quality Handbook ANNEX: Sprat in the North Sea
Stock specific documentation of standard assessment procedures used by ICES.

| Stock: | Sprat in the North Sea |
| :--- | :--- |
| Working Group | Herring Assessment Working Group (HAWG) |
| Date: | $22^{\mathrm{TH}}$ March 2009 |
| Authors | E. Torstensen, L. W. Clausen, C. Frisk, C. Kvamme. |

## A. General

## A.1. Stock definition

Sprat in ICES area IV.

## A.2. Fishery

The Danish small meshed fishery is responsible for the majority of the landings. A study undertaken in 2000 showed that the species composition in the Danish sprat fishery has changed towards a fishery with low by-catches of other species (ICES CM 2001/ACFM:12). The Norwegian sprat fishery is mainly carried out by purse-seiners. A closure of the Norwegian fishery was introduced for the second and third quarter in 1999 and this management regime is still in force. On top of this management regime, a maximum vessel quota is set for the Norwegian vessels; and they are not allowed to fish in the Norwegian Economic Zone until the Norwegian quota in EU waters has been taken.

## A.3. Ecosystem aspects

Multispecies investigations have demonstrated that sprat is one of the important prey species in the North Sea ecosystem. Many of the plankton-feeding fish species have recruited poorly in recent years (e.g. herring, sandeel, Norway pout) possibly due to changing availability of prey. The influence of the sprat fishery for other fish species sea birds and sea mammals, are at present unknown.

Sprat is an important part of the pelagic ecosystem in the North Sea. As plankton feeders, sprat is an essential prey species for higher trophic levels. Both as juveniles and as adults they are an important source of food for other fish, sea birds and mammals.

The zooplankton community structure that is sustaining the sprat stock appears to be changing, and there has been a long-term decrease in total zooplankton abundance in the northern North Sea (Reid et al., 2003; Beaugrand, 2003; ICES, 2006). However, sprat is mainly distributed in the southern North Sea where these trends have not been observed (ICES, 2006). The implications of the environmental change for sprat are unknown.

Sprat spawns in the upper water layers. In the North Sea sprat eggs and larvae are found more or less during the whole year. Spawning and nursery areas, being near the coast, are particularly sensitive and vulnerable to anthropogenic influences.

## B. Data

## B.1. Commercial catch

Commercial catch data is provided by the national laboratories of nations exploiting sprat in the North Sea. The labs have used a spreadsheet to provide landings and sampling data. The sampling intensity for biological samples, i.e., age and weight-atage is mainly performed following the EU regulation 1639/2001 as the country landing most of the catches (Denmark) follows this regulation. This provision requires 1 sample per 2000 tonnes landed. This sampling level is lower than the guidelines (1 sample per 1000 tonnes) previously used by the HAWG. As the fishery is carried out in a limited area, the recommended sampling level can be regarded as adequate.

The majority of commercial catch and sampling data are submitted in the Exchange sheet v. 1.6.4. This method is now run in parallel with INTERCATCH, which is maintained by ICES. INTERCATCH is still in development and is not completely satisfactory in terms of flexibility and outputs. Thus HAWG uses both. The data in the exchange spreadsheets are samples allocated to catch using the SALLOCLapplication (Patterson, 1998). This application gives the needed standard outputs on sampling status and biological parameters. It also clearly documents any decisions made by the stock co-ordinators for filling in missing data and raising the catch information of one nation/quarter/area with information from another data set. The stock co-ordinator allocates samples of catch numbers, mean length and mean weight-at-age to unsampled catches using appropriate samples by gear (fleet), area and quarter. If an exact match is not available then a neighbouring area in the same quarter is used.

## B.2. Biological

Mean weights at age in the catch in the $1^{\text {st }}$ quarter are used as stock weights.
Natural mortality: Results from the North Sea multi-species VPA (ICES CM 2002/D:04) can be used as an estimate of the predation mortality. To estimate total natural mortality a value of 0.2 to account for other sources of natural mortality should be added to the predation mortality.

## B.3. Surveys

The acoustic surveys for the North Sea herring in June-July have estimated sprat abundance since 1996 (ICES 2009/LRC:02) (see table in section 8.3.2). In this period no sprat has been recorded in the northern part of the North Sea. The sprat has almost exclusively been found in the eastern and southern parts of the North Sea (Figure 8.3.2). The age-disaggregated time series of sprat abundance and biomass from the acoustic series (ICES areas IVa-c), have been re-calculated using FishFrame for the years 2003-2007 (ICES CM 2008/LRC:01). The surveyed area has increased over the years, thus only figures for the last 6 years are roughly comparable.

The acoustic recordings are carried out using a Simrad EK60 38 kHz sounder echo-integrator with transducers mounted on the hull, drop keel or towed bodies. Prior to 2006, Simrad EK500 and EY500 were also used. Further data analysis is carried out using either BI500, Echoview, Echoann software or LSSS. The survey track is
selected to cover the area giving a basic sampling intensity over the whole area based on the limits of herring densities found in previous years. A transect spacing of 15 nautical miles is used in most parts of the area. The surveys are co-ordinated under ICES Planning Group for International Pelagic Surveys (PGIPS, former PGHERS).

The IBTS (February) sprat indices (no. per hour) in Div. IV were previously used as an index of abundance of sprat in the North Sea. The fishing gear used in the IBTSsurvey was standardised in 1983. The complete time series of the IBTS Q1 index, from 1984 onwards, is calculated by the ICES DATARAS database (http://datras.ices.dk). The index is calculated as a weighted mean (over all squares sampled in a particular year) of the mean CPUE in a stat-square. Symbolically, this can be represented as:

$$
I=\frac{\sum_{s s} \omega_{s s} \frac{\sum_{i}^{n_{s s}} C P U E_{s s, i}}{n_{s s}}}{\sum_{s s} \omega_{s s}}
$$

where $I$ is the index value, $\omega_{s s}$ is the weighting factor given to an individual stat square, $n_{s s}$ is the number of hauls in a given stat square, and CPUEss,i is the CPUE of haul $i$ in a given stat-square. Only stat-squares that are sampled in the particular year are included the summations. The weighting factors for each stat-square are proportional to the volume of water in a square within a certain water depth. The limits of this water depth are uncertain and have been reported earlier as 10-150 m (ICES 1995/Assess: 13) and as 20-100 m (pers. comm. ICES staff, 2009). The weighting factors are given in the following table.

| Area | Weight | Area | Weight | Area | Weight | Area | Weight | Area | Weight | Area | Weight | Area | Weight | Area | Weight |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 31F1 | 0.6 | $36 F 3$ |  | 38F5 | 1 | 40F5 |  | 42F3 |  | 44F2 |  | 47E8 |  | 50 E 7 | 0.6 |
| 3172 | 0.8 | 3674 | 1 | 3856 | 1 | 40F6 | 1 | 42 F 4 | 1 | 4473 | 1 | 47E9 | 1 | 50 E 8 | 0.7 |
| 32 F 1 | 0.8 | $36 F 5$ | 1 | $38 \mathrm{F7}$ | 1 | 40F7 | 1 | 42 F 5 | 1 | 4474 | 1 | 47F0 | 1 | 50E9 | 0.9 |
| 32 F 2 | 1 | $36 F 6$ | 0.9 | 38F8 | 0.3 | 40F8 | 0.1 | 42 F 6 | 1 | 44 F 5 | 0.9 | 47F1 | 1 | 50FO | 1 |
| 32 F 3 | 0.8 | $36 F 7$ | 0.4 | 39 E 8 | 0.5 | $41 \mathrm{E6}$ | 0.03 | $42 \mathrm{F7}$ | 1 | 45E6 | 0.4 | 47F2 | 1 | 50F1 | 1 |
| 32F4 | 0.01 | 36F8 | 0.5 | $39 \mathrm{E9}$ | 1 | $41 \mathrm{E7}$ | 0.8 | 42F8 | 0.2 | 45E7 |  | 47 F 3 | 0.6 | 50F2 | 1 |
| 33 F 1 | 0.3 | $37 \mathrm{E9}$ | 0.2 | 39F0 | 1 | 41 E 8 | 1 | $43 \mathrm{E7}$ | 0.03 | 45E8 |  | 48E6 | 1 | 50F3 | 0.2 |
| $33 F 2$ | 1 | 37 FO | 1 | 3971 | 1 | $41 \mathrm{E9}$ | 1 | 43 E 8 | 0.9 | 45E9 | 1 | 48 E 7 | 1 | $51 \mathrm{E6}$ | 0 |
| 33 F 3 | 1 | 37 F 1 | 1 | $39 F 2$ | 1 | 41F0 | 1 | $43 \mathrm{E9}$ | 1 | 45FO |  | 48 E 8 | 0.9 | $51 \mathrm{E7}$ | 0 |
| 3374 | 0.4 | 37 F 2 | 1 | 3973 | 1 | 41F1 | 1 | 43F0 | 1 | 45 F 1 |  | 48E9 |  | 51 E 8 | 0.5 |
| 3471 | 0.4 | 37 F 3 | 1 | 3974 | 1 | 41F2 | 1 | $43 F 1$ | 1 | 45F2 | 1 | 48 FO |  | $51 \mathrm{E9}$ | 1 |
| 34F2 | 1 | 3774 | 1 | 3975 | 1 | 41F3 | 1 | $43 F 2$ | 1 | 45 F 3 | 1 | 48 F 1 | 1 | 51 F0 | 1 |
| 3473 | 1 | 3775 | 1 | 3976 | 1 | 41F4 | 1 | $43 F 3$ | 1 | $45 F 4$ | 0.6 | 48 F 2 | 1 | $51 F 1$ | 1 |
| 3474 | 0.6 | $37 \mathrm{F6}$ | 1 | 3977 | 1 | 41F5 | 1 | $43 F 4$ | 1 | 46E6 | 0.4 | 48 F 3 | 0.5 | 51 F 2 | 0.5 |
| 35FO | 0.8 | $37 F 7$ | 1 | 3978 | 0.4 | 41F6 | 1 | $43 F 5$ | 1 | $46 E 7$ | 0.9 | 49E6 | 0.8 | 51F3 | 0 |
| 3571 | 1 | 3778 | 0.8 | 40E7 | 0.04 | $41 F 7$ | 1 | $43 F 6$ | 1 | 46 E8 |  | $49 E 7$ | 1 | 52 E 6 | 0 |
| 35F2 | 1 | 38 E 8 | 0.2 | 40E8 | 0.8 | 41F8 | 0.1 | $43 F 7$ | 1 | 46E9 |  | 49 E 8 | 0.4 | 52 E 7 | 0 |
| 35F3 | 1 | $38 \mathrm{E9}$ | 0.9 | 40E9 |  | $42 \mathrm{E7}$ | 0.4 | 44E6 | 0.5 | 46F0 |  | 49 E 9 | 1 | 52 E 8 | 0 |
| 3574 | 0.9 | 38 FO | 1 | 40FO | 1 | 42 E 8 | 1 | $44 \mathrm{E7}$ | 0.5 | 46F1 | 1 | 49F0 |  | 52 E 9 | 0.1 |
| 3575 | 0.1 | 38 F 1 | 1 | 40F1 |  | $42 \mathrm{E9}$ |  | 44E8 | 0.9 | 46F2 | 1 | 49F1 | 1 | 52F0 | 0.2 |
| 36F0 | 0.9 | 38 F 2 |  | 40F2 |  | 42FO | 1 | 44E9 | 1 | 46F3 | 0.8 | 49F2 | , | 52 F 1 | 0.5 |
| 36F1 | 1 | 38 F 3 | 1 | 40F3 | 1 | 42 F 1 | 1 | 44FO | 1 | $47 \mathrm{E6}$ | 0.8 | 4973 | 0.5 | 52 F 2 | 0.1 |
| 36F2 |  | 38F4 |  | 40F4 |  | 42 F 2 | 1 | 44F1 |  | 47E7 | 0.6 | 50 E 6 | 0.1 | 52F3 | 0 |

## B.4. Commercial CPUE

Not used for this stock.

## B.5. Other relevant data

None

## C. Historical Stock Development

Sprat is a relatively short-lived species. The stock and the catches are consisting mostly of 1 and 2 year-olds. In addition, there are difficulties in age reading resulting in unreliable estimates of numbers at age both from the surveys and the commercial catch. Given those limitations a data exploration using Catch-Survey Analysis (CSA) was undertaken by the WG from 2003 to 2008 as an exploratory assessment (ICES CM 2008/ACOM:02). CSA is an assessment method designed for cases where full agestructured data are missing.

## D. Short-Term Projection

Not performed.

## E. Medium-Term Projections

Not performed

## F. Long-Term Projections

Not performed

## G. Biological Reference Points

Not set.

## H. Other Issues

None

## I. References

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## Annex 10 - Stock Annex Sprat in Division IIla

Quality Handbook ANNEX: Sprat IIIa
Stock specific documentation of standard assessment procedures used by ICES.

| Stock: | Sprat in Division IIIa |
| :--- | :--- |
| Working Group: | Herring Assessment Working Group (HAWG) |
| Date: | 22th March 2009 |
| Authors: | Torstensen, E.; Clausen, L.W., Frisk, C., Kvamme, C. |

## A. General

## A.1. Stock definition

Sprat in ICES area IIIa

## A.2. Fishery

Fleets from Denmark, Norway and Sweden carry out the sprat fishery in Division IIIa. The Danish sprat fishery consists of trawlers using a < 32 mm mesh size and the landings are used for fishmeal and oil production. Some of the sprat landings from Denmark and Sweden are by-catches in the herring fishery using 32 mm mesh-size cod ends. The Swedish fishery is directed at herring with by-catches of sprat. The Swedish fleet is mainly pelagic trawlers and also a few purse seiners. The Norwegian sprat fishery in Division IIIa is an inshore purse seine fishery (vessels $<27.5 \mathrm{~m}$ ) for human consumption.

The majority of the landings are made by the Danish fleet. In 1997 a mixed-clupeoid fishery management regime was changed to a new agreement between the EU and Norway that resulted in a TAC for sprat as well as a by-catch ceiling for herring. Catches are taken in all quarters, though with the bulk of catches in the first and fourth quarter. Denmark has a total ban on the sprat fishery in Division IIIa from May to September.

There was a considerable increase in landings from about 10,000 t in 1993 to a peak of $96,000 \mathrm{t}$ in 1994. From 1996 the landings have varied between 9,000 t (2008) and 40,000t (2005).

## A.3. Ecosystem aspects

No information of the ecosystem and the accompanying considerations are known at present. In the adjacent North Sea Multispecies investigations have demonstrated that sprat is one of the important prey species in the North Sea ecosystem, as a prey species for both fish and seabirds. Many of the plankton feeding fish have recruited poorly in recent years (eg. herring, sandeel, Norway pout). The implications for sprat in IIIa are at present unknown.

## B. Data

## B.1. Commercial catch

Commercial catch data are submitted to ICES from the national laboratories belonging to nations exploiting the sprat in Division IIIa. The sampling intensity for biological samples, i.e., age and weight-at-age is mainly performed following the EU regulation 1639/2001 as Denmark, landing most of the catches, follows this regulation. This provision requires 1 sample per 2000 tonnes landed.

The majority of commercial catch and sampling data are submitted in the Exchange sheet v . 1.6.4. This method is now run in parallel with INTERCATCH, which is maintained by ICES. INTERCATCH is still in development and is not completely satisfactory in terms of flexibility and outputs. Thus HAWG uses both. The data in the exchange spreadsheets are samples allocated to catch using the SALLOCLapplication (Patterson, 1998). This application gives the needed standard outputs on sampling status and biological parameters. It also clearly documents any decisions made by the stock co-ordinators for filling in missing data and raising the catch information of one nation/quarter/area with information from another data set.

The stock co-ordinator allocates samples of catch numbers, mean length and mean weight-at-age to unsampled catches using appropriate samples by gear (fleet), area and quarter. If an exact match is not available then a neighbouring area in the same quarter is used.

## B.2. Biological

Mean-weight-at-age for all ages is in the same order as the previous years, except for 2007 where the mean weight-at-age were the largest in the period. Mean weights-atage for 1996-2003 are presented in ICES CM 2005/ACFM:16.

No estimation of natural mortality is made for this stock.

## B.3. Surveys

Acoustic estimates of sprat have been available from the ICES co-ordinated Herring Acoustic surveys since 1996. The estimated biomass of sprat has been very variable with low values in the period from 1997 to 2002, but recently the biomass has increased. In Division IIIa, sprat has mainly been observed in the Kattegat.

The IBTS (February) sprat indices (no per hour) in Division IIIa have been used as an index of abundance. In later years, the index has not been considered useful for management of sprat in Division IIIa. The indices are calculated as mean no./hr (CPUE) weighted by area where water depths are between 10 and 150 m (ICES 1995/Assess:13). The indices were revised in 2002 (ICES 2002/ACFM:12) based on an agreement in the IBTS WG in 1999, where it was decided to calculate the sprat index as an area weighted mean over means by rectangles for the IIIa (ICES 1999/D:2). The old time-series of IBTS indices (from 1984-2001) is shown in ICES 2001/ACFM:10.

## B.4. Commercial CPUE

Not used for this stock.

## B.5. Other relevant data

None

## C. Historical Stock Development

Not performed

## D. Short-Term Projection

Not performed

## E. Medium-Term Projections

Not performed

## F. Long-Term Projections

Not performed

## G. Biological Reference Points

Not set
H. Other Issues

None

## I. References

ICES 1995. Report of the Herring Assessment Working Group for the Area South of $62^{\circ} \mathrm{N}$. ICES 1995/Assess:13

ICES 1999. International Bottom Trawl Survey in the North Sea, Skagerrak and Kattegat in 1998. ICES 1999/D:2

ICES 2001. Report of the Study Group on the Herring Assessment Units in the Baltic Sea. ICES CM 2001/ACFM:10.

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ICES 2005. Herring Assessment Working Group for the Area South of $62^{\circ} \mathrm{N}$ (HAWG). ICES CM 2005/ACFM:16.

Patterson, K.R. 1998: A programme for calculating total international catch-at-age and weight-at-age. Working Document to Herring Assessment Working Group South of $62^{\circ} \mathrm{N}$. ICES CM 1998/ACFM:14.

## Annex 11 Stock Annex - Sprat in Division VIIde

Quality Handbook ANNEX:_Sprat VIIde
Stock specific documentation of standard assessment procedures used by ICES.

| Stock: | Sprat in Division VIId,e |
| :--- | :--- |
| Working Group: | Herring Assessment Working Group (HAWG) |
| Date: | $25^{\mathrm{TH}}$ March 2009 |
| Author: | Torstensen, E; Clausen, L.W., Kvamme, C. |

A. General
A.1. Stock definition

Sprat in ICES area VIId, VIIe

## A.2. Fishery

Vessels from UK (England and Wales) are currently responsible for the catches. The landings in this area are small and have never been above $6,000 \mathrm{t}$ since 1985. Since 2000 the landings have been in the range of 840 t (2004) and 3370 t (2008)

## A.3. Ecosystem aspects

None
B. Data

## B.1. Commercial catch

The commercial catch is provided by the national laboratories belonging to nations exploiting the sprat in the Division VIId and VIIe.

## B.2. Biological

Sampling for biological samples, i.e. age and weight-at-age has not been performed since 1999, but as the fishery is so small, this is not considered to be a problem.

## B.3. Surveys

There are no surveys targeting sprat in this area.

## B.4. Commercial CPUE

Not used for this stock.

## B.5. Other relevant data

None

## C. Historical Stock Development

Not performed

## D. Short-Term Projection

Not performed

## E. Medium-Term Projections

Not performed
F. Long-Term Projections

Not performed
G. Biological Reference Points

Not set
H. Other Issues

None
I. References

## Annex 12 - Celtic Sea Herring

## Benchmark Review

6 May 2009 by correspondence
Reviewers:
Steve Cadrin, USA (chair)
Lionel Pawlowski, France
...with assistance from Daniel Goethel and Lisa Kerr, USA
"This annex describes the work done on Celtic Sea Herring immediately after the HAWG. The benchmark performed at the HAWG 2009 is reviewed (section 1). The management plan proposed for the stock has been evaluated (section 2) and consequently reviewed (section 3)."

## Summary

Methodology of the Celtic Sea Herring stock assessment was reviewed according to the stock annex (Annex 05 Celtic Sea and Division VIIj_V1_09.doc), and the most recent application of the methodology was reviewed according to the 2009 stock assessment (04-Celtic Sea and Div. VIIj_Herring_2009.doc). Our review addresses the following questions from the ICES Secretariat (B. Schoute 22 April 2009):

1. Is the new assessment methodology correct, of high standard and does it make optimal use of the available data?
2. Are the settings of the forecast (only short term, medium and long term are not relevant here) chosen correctly and do the reference points still apply.
3. Is the methodology adequately described in the stock annex, meaning that the assessment can in principle be carried out by experienced outsiders on the basis of this text.

1 ) We conclude that the new assessment methodology is generally sound, but some inconsistencies between surveys remain, results are somewhat uncertain, and assumptions of the model should be further explored. The assessment results are relatively consistent, but calibration relationships are weak and based on a short survey time series. Among the alternative assessment models explored, the revised configuration of Integrated Catch Analysis (FLICA) appears to be most appropriate model of data from the fishery and resource. However, we encourage the assessment Working Group to continue to explore more advanced stock assessment models to make optimal use of the available data. The following sources of uncertainty should be addressed in subsequent assessments:
a ) The ICA calibration is based on a short survey series, and the calibration relationships are weak, with some year effects. The resulting estimates of terminal SSB are imprecise, and much of the uncertainty in terminal SSB is not included in estimates of precision. For example, a large portion of the spawning stock is composed of 1-ringers, for which proportion mature is poorly understood, and geometric mean abundance is assumed in the terminal year.
b) A stock assessment model that relaxes the assumption of constant selectivity in recent years should be explored. The separability assumption (and the assumed selectivity at the oldest ages for the entire time series) may lead to misinterpretation of the apparent shifts in age selectivity by the fishery. According to the 2009 HAWG report (page 298), the 2007 benchmark concluded that changes in fishing pattern (and conflicting signals) prevented a final assessment from being conducted
c) The magnitude of discarded catch should be estimated and included in the stock assessment.
d) A stock assessment approach that accounts for the mixed-stock resource and connectivity with adjacent management units should be developed.
e) Fishing mortality reference points $\left(\mathrm{F}_{\lim }\right.$ and $\mathrm{F}_{\mathrm{pa}}$ ) should be proposed.
f) Consumption of Celtic Sea herring should be estimated and considered for stock assessment and fishery management.
2 ) The settings of the short-term forecast appear to be chosen correctly, and the $B_{l i m}$ and $B_{p a}$ reference points (as revised for retrospective change) are still appropriate. However, we suggest that the same forecast approach be extended for long-term, stochastic projection to determine the fishing mortality rate associated with $B_{\lim }$ (as a candidate for $\mathrm{F}_{\text {lim }}$ ) and its uncertainty (to derive $\mathrm{F}_{\mathrm{pa}}$ and potentially a revised $\mathrm{B}_{\mathrm{pa}}$ ) as well as MSY reference points.
3 ) The methodology is generally well-described in the stock annex and allows repeatability. The various tables and information in the body of the report give the strong impression that all inputs are sufficiently documented to allow an outsider to do an assessment. However, some details of the most recent application of the stock assessment model (e.g., input and output tables, model diagnostics) should be provided to justify the modeling decisions.

## Detailed Comments (organized by Annex section)

## A. General

A.1. Stock definition - Several aspects of stock definition are described to justify the appropriateness of the management unit and identify aspects of population structure that may influence stock assessment and fishery management. Atlantic herring are 'population-rich' throughout their range, with complicated patterns of ontogenetic movement and mixing of spawning groups. These complex patterns present challenges to conventional stock assessment and fishery management.

1 ) Combined assessment of autumn and winter spawning groups appears to be the most appropriate use of available data, because of extensive mixing of spawning groups resulting in mixed-group fisheries and surveys. However, continued advancements in discrimination of seasonal spawning groups should be explored with the ultimate goal of stock composition analysis and consideration of spawning groups in assessment and management.
2 ) The inclusion of area VIIj in the Celtic Sea management unit appears to be appropriate because of similar demographic patterns in VIIj and g, larval mixing between the two areas, and a common nursery area in VIIj shared by herring spawned in VIIg. However, Figure 1 suggests that spawning in

VIIg is primarily in winter, while spawning in VIIj occurs in both autumn and winter. Similar to the comment above, the seasonal spawning pattern suggests that the development of stock composition analysis would facilitate the consideration of spawning groups in assessment and management.
3 ) The boundary between the Celtic Sea and Irish Sea herring management units is supported by the results of an extensive multidisciplinary program (WESTHER). Although the $52^{\circ} 30^{\prime} \mathrm{N}$ boundary is well-justified, advection of larvae from the Celtic Sea to the Irish Sea and subsequent return to spawn in the Celtic Sea has consequences to assessment and management of both resources. Loss of larvae from the Celtic Sea will add noise to the stock-recruit relationship. Depending on which age fish return to the Celtic Sea, the immigration may confound inferences of mortality from the catch-at-age analysis which assumes a closed population. Return migration at 1-ringers will have less influence on the perceived population dynamics than on immigration of older ages. The relationship between Celtic Sea and Irish Sea herring should be further investigated to better understand the sensitivity of the closed-population assumption in the assessment.
4 ) Similar to the comment above, the relationship between Celtic Sea herring and those in in VIIe-f and VIIIa should be investigated to understand the sensitivity of the closed-population assumption in the assessment.
5 ) Figures 1 and 2 are switched.
6 ) A Figure of the region should be provided that includes all of the areas described in the Annex (VIIe, VIIh, VIIk, VIIIa).
7 ) The species name Clupea harengus should be included in the Annex.

## A.2. Fishery

1 ) The fishery description is informative and well-written.
2 ) The increased landings after World War II support the premise that fishing influences stock size - a principle that should not be taken for granted for small pelagic species.
3 ) Any information on historical landings (prior to 1958) would be informative.
4 ) The statement that "Further fluctuations in the landings were evident during this time with high quantities of herring landed from 1958-1960 and from 1966-1971 (Molloy, 1972)" is somewhat inconsistent with the data plotted in Figure 6, in which annual landings from 1959 to 1963 are similar (i.e., the 'high quantities from 1958-1960 persisted to 1963).

5 ) The 'polyvalent' category of vessels should be described as in the 2009 HAWG report ("The term 'Polyvalent' refers to a segment of the Irish fleet, entitled to fish for any species to catch a variety of species, under Irish law" page 296).
6 ) The catch of large, old fish appears to depend on what areas are seasonally open to the fleet (e.g., Labadie Bank being open in July 2003 led to older fish being caught as compared to openings later in the year). Therefore, two aspects of the assessment model (separability in the recent period and full selectivity of the oldest age in all years) may be inappropriate.
7 ) The last paragraph "the Irish Quota" is redundant almost word for word with the second paragraph of the "Fishery in recent years"

## A.3. Ecosystem aspects

1 ) Given the important role of herring in the ecosystem, more information is needed on consumption of Celtic Sea herring and predation of other species by herring. Although estimates of herring consumption in the North Sea are used to derive the assumed natural mortality rates for Celtic Sea herring, consumption of herring in the North Sea should be estimated, and incorporation of consumption in stock assessment and management should be considered.
2 ) References should be provided for the statement that "studies have been unable to demonstrate that changes in the environmental regime in the Celtic Sea have had any effect on productivity of this stock."
3 ) The recent reduction in size-at-age should be reported in this section as a possible response to ecosystem factors. The trend is critical for estimation of spawning biomass, and the cause of the trend is important for assessment decisions and modeling future expectations (e.g., forecasts and reference points).
4 ) Similarly, the increasing recent trend in total mortality estimates from catch curves or $\log$ catch ratios and the decreasing recent trend in fishing mortality from the stock assessment model suggest an increase in natural mortality, which may reflect ecosystem change.
5 ) The ecosystem description and the summary of spawning dynamics suggest that there is adequate information to develop a bio-physical model of larval transport that would provide a complementary perspective on connectivity among spawning groups.
6 ) Differences in survival between the Irish and Celtic Seas could have important consequences to population dynamics. The relative contribution of each habitat to the adult population in the Celtic Sea should be determined.
7 ) Given that the Celtic Sea is near the southern extent of the range of herring, and the increasing trend in temperature, the potential for a northward shift in distribution should be monitored.

## Discards

1 ) The "discards" section of this part should probably be in the data section.
2 ) Although the discard rates are considered to be low and discards are not included in the assessment, discards occur. The report indicates that discarding is influenced by market situations, which suggests it may rise. While, it does not seem to have at the moment a potential impact on the assessment, it could be useful to evaluate the consequence of the inclusion of discards on an exploratory basis. The underestimate of total catch produces biased estimates of stock size and mortality. The discard rate estimated by Berrow, et al. (1998), $4.7 \%$, should be used to derive an approximate magnitude of discards that would be more accurate than the implicit assumption of no discards.
3 ) An at-sea monitoring program should be developed to estimate discard rates (including slippage) and to sample size and age structure of discards.
4 ) The statement that Berrow, et al. (1998) "indicated that the Celtic Sea herring fishery is very selective and that discard rates are well within the figures estimated for fishery models" is not clear. Does the statement suggest
that a $4.7 \%$ discard rate is similar to that estimated for other fisheries, or similar to the rate assumed in fishery models? Celtic Sea herring stock assessment assumes no discards and is not consistent with the estimate of a $4.7 \%$ discard rate.

## B. Data

## B.1. Commercial catch

1) Sampling intensity of the series of catch-at-age should be provided to evaluate the reliability of catch-at-age estimates. For example, Table 4.2.2.1 in the 2009 HAWG report indicates that 45 samples were collected from the 2008 fishery, and all major area-quarter components of the catch were sampled. Is this typical of the sampling intensity since 1958 or are there any systematic gaps in historical sampling that should be considered in the interpretation of catch-at-age? Given the complex pattern of time-area closures and fishing patterns, 45 samples per year may not adequately characterize some of the fine-scale differences in catch-at-age.
2 ) The report mentions the landings statistics for this stock need correction for misreporting. Landings apparently include substantial amounts ( $>10 \%$ ) of fish from other areas but while the correction is made (i.e. unallocated landings), no information is given on how this correction is done. This information is important to explain how to prepare the data from the raw landings statistics.
3 ) A requirement of logbook data for all vessels in the sentinel fishery could improve estimates of small boat landings.

## B.2. Biological

1 ) The various biological parameters are well described and their quality appears to be reliable for the assessment.
2 ) Have the age determination methods been validated?
3 ) Including some typical age-length keys would be helpful to evaluate how well catch-at-age is being estimated, particularly at older ages.
4 ) Including the quality-control results for precision estimates would be informative, and potentially useful for the development of advanced statistical catch-at-age models that use the pattern of disagreements to model errors in the catch-at-age.
5 ) The use of age-specific natural mortality rates from multispecies VPA is appropriate for a small, pelagic forage species, but a development of a MSVPA for the Celtic Sea would be more appropriate than using the results from the North Sea. A reference should be provided for the MSVPA so that its details do not need to be included in the Annex.

6 ) The cause of the reduction in weight-at-age should be explored further. More specifically, determining if it results from ecosystem factors or fishing patterns is essential for making the correct selectivity assumptions in the stock assessment. It would be valuable to inspect weight-at-age data from surveys to see if fishery-independent data reflect the same recent reduction. The beginning of the decline is consistent with the development of the roe fishery.
7 ) The choice of the maturity ogive suggests that various sources of information provide similar results. However, the amount of available information
(number of individuals sampled) is not included. The rationale for assuming $50 \%$ maturity of 1 -ringers, rather than the estimated $58 \%$, is that the fishery probably samples precocious fish. However, the estimate of $58 \%$ is from a survey. The text does not report if the estimate of $>50 \%$ by Lynch (in prep.) is from the fishery or a survey. Given the substantial contribution of 1-ringers to the spawning stock, a more precise estimate of proportion mature should be applied.
8 ) Recruitment from the Irish sea may affect maturity of the population (and consequently maturity ogives). The possible influence of individuals from the Irish Sea raises the question of the proportion of individuals from that area and the effects of possible changes in maturity from one area to another. For some other stocks, like the Celtic Sea cod, a similar situation is observed and the lack of samples makes any maturity ogive rather uncertain. This was one of the criticisms in the benchmark review of the Celtic Sea cod.

## B.3. Surveys

1) The timing of the survey appears to be related to 'year effects' in calibration diagnostics (HAWG 2009 Figure 4.6.1.1), suggesting that the portion of the resource in the survey area is sensitive to the time of the survey. The 2002/2003 survey (conducted in September and October) has all positive residuals (i.e., more fish in the survey than predicted by the model), and the 2003/2004 survey (conducted in October and November) has all negative residuals (i.e., fewer fish in the survey than predicted by the model), suggesting that fewer fish are available to the survey later in the year. Is it possible that spawners are in the process of leaving the survey area during the survey?
2 ) Imprecision of survey estimates is illustrated by the large difference between estimates of 2008 SSB from the survey ( 90 kt ) and the assessment model (56kt).
3 ) For the acoustic survey, the estimates of CV appear to be based on a simple function of the positive number of samples. Some explanations about that relationship would have been welcome. An apparent contradiction is in the text: "CV was obtained based on transect mean densities," but mean density is not included in the equation. There appears to be something missing from the equation.
4 ) The decision to use a shorter, standardized series for a tuning index is valid.
5 ) More information is needed to describe how indices of abundance at age are derived from the acoustic survey.
6 ) Tuning is based only on the acoustic survey which apparently provides the best indices. Data from other surveys are not used. Some other stocks, like the Celtic Sea cod, have the same issues of having surveys that are not specifically targeting those species sampling few and variable numbers of fish. Some work has been carried out during the WKROUND benchmark to combine survey indices and some others stocks (e.g. Sole in VIId) use combined survey indices. The report and stock annex do not mention any attempt to use or combine the available information. Some exploratory work on using those datasets would also be welcome. This could involve evaluating how the indices behave against each other, against fishing vessels
and how they could affect the assessment. Some analysis of the trends of all survey data would be helpful to support the choice of only using the acoustic survey for the assessment.
7 ) The analysis of productivity over time is sufficiently commented to naturally end with the conclusion that recent F has been detrimental to the stock productivity.
8 ) If the Irish Groundfish Survey is expected to provide qualitative information for the assessment, results should be included in the Annex.
9 ) Similarly, if the Northern Ireland GFS survey offers a potential recruitment index, more details are needed in the Annex so that it can be considered as more information becomes available on natal origin.
10 ) Similarly, data from larval surveys should be provided as a comparison to stock assessment results.
B.4. Commercial CPUE - The decision to exclude fishery CPUE as a tuning index in the stock assessment is valid, because of the nature of herring behavior, fishing patterns and management changes. However, it would be informative to compare the acoustic survey index to CPUE information from the fishing vessels. Some stocks (e.g. whiting in the North Sea) have conflicting patterns between surveys and fishing vessels. One reason could be some slight changes in the survey interfering with the results. Therefore, this type of comparison can be helpful to evaluate the consistency of the observation from the surveys in addition to the quality (i.e. level of noise) of the data which is another aspect to consider.

## C. Historical Stock Development

1 ) This section shows some issues with noisy data but does not seem to explain the "conflicting signals in input data and changes in the fishing pattern" referenced in the HAWG report from the 2007 benchmark.
2 ) The description of 'time periods in the fishery' is informative and suggests that some of the selectivity assumptions in the stock assessment should be reconsidered. The roe fishery targeted older, mature fish, which would lead to greater selectivity of the oldest age during that period.
3 ) Tables 3 and 4 referenced in 'Time Periods in the Fishery' should be Tables 4 and 5.
4 ) Estimates of total mortality from log catch ratios and catch surveys are informative, but the age ranges selected for catch-curve analysis are inconsistent with results from the stock assessment model, because of the assumed pattern of natural mortality at age and estimated selectivity at age. According to the assessment model, herring are not fully selected until age-3, and natural mortality of ages 2 and 3 is greater than than for ages $4+$, so catch curves should be revised from ages 2-7 to ages 4-7.
5 ) The increasing recent trend in total mortality estimates from catch curves or log catch ratios appear to contradict the decreasing recent trend in fishing mortality from the stock assessment model. The cause of the discrepancy (e.g., increasing natural mortality) should be explored.
6 ) In the assessment section, the HAWG report (4.6.1) mentions "conflicting signals in input data and changes in the fishing pattern" but no information is given on the "historic" choices made for exploratory assessments. The decisions about model configuration are well explained and the assessment well commented (changes in plus groups, shortening time series,
terminal selection and reducing the separable period) providing the rationale for the parameters used in the final assessment.
7 ) The concern raised by the 2007 benchmark assessment about violating the assumed constant separable pattern was not addressed. The assessment model still assumes separability in recent years. We reiterate the concern about assuming constant selectivity and repeat the recommendation to consider alternative modeling approaches that relax this assumption.

8 ) The revision of the catch-at-age used for the stock assessment model (truncation to ages $1-6+$ ) produces a more realistic selectivity pattern than the previous analysis of ages 1-7+.
9 ) Diagnostic features of the stock assessment model are needed in the Annex to evaluate model performance. Standard diagnostics from the 2009 assessment should be included (e.g., HAWG 2009 Figures 4.6.1.2-3 model residuals, 4.6.1.4-5 confidence intervals, 4.6.1.6. historical comparisons, 4.6.2.1 calibration plots, 4.6.2.5-6 retrospective analysis). Inspection of diagnostics suggests that the ICA model is relatively consistent and has no strong patterns in catch residuals, but there are strong 'year effects' in the survey residuals (i.e., same direction of deviation at all ages) for the first three years of the six surveys used (Figure 4.6.1.1), and calibration relationships are relatively weak (Figure 4.6.2.1). A 'year effect' in the terminal year will present problems for estimating terminal abundance and determining stock status.
10 ) The advantage of ICA over other models is well explained as well as the reasons for adopting new parameters. Considering the few changes in the list of parameters, it can be confusing to have the parameters listed for the former and new assessment methods separately. Maybe combining both sets of parameters into a single table would be more useful (considering only 2 parameters of 8 changed) so no "quick reader" may switch to the wrong set of parameters.
11 ) The Annex reports that ICA was chosen because of its emphasis on young ages and greater consistency, but there is no information in the Annex or the 2009 HAWG report on the performance of alternative models. It is difficult to judge the validity of that conclusion without example results from all viable models. For example, did XSA also have year effects in survey residuals?

12 ) The ICA model appears to perform well for this application, but the method is somewhat dated (it is a re-codification of the CAGEAN model developed by Deriso et al. 1985). Catch-at-age models have evolved since the 1980s, and more advanced methods (e.g., statistical catch-at-age, SCAA) may be more appropriate for assessing the data available for Celtic Sea herring. SCAA would also be able to use all recent and historical information available (e.g., selectivity for each fishery and each period, calibration of historical abundance with discontinued surveys, discard rate estimates)

13 ) References for ICA (Patterson 1998) and FLICA (flr-project.org) are needed.
14 ) The analysis of productivity over time supports the conclusion that recent F has been detrimental to the stock productivity. However, the calculation of surplus production is either poorly described or inaccurate. The equa-
tion $\mathrm{Ps}=\mathrm{Br}+\mathrm{Bg}-\mathrm{M}$ does not account for the different units of biomass $(\mathrm{t})$ and natural mortality rate $\left(\mathrm{y}^{-1}\right)$. The inappropriate mix of instantaneous rates and biomass is continued in the subsequent statement that net production is calculated as Ps - F. Surplus production should be calculated as $\mathrm{Ps}=\mathrm{Br}+\mathrm{Bg}-\mathrm{Bm}$ where Bm is biomass of fish that die of natural causes. Total Production should be calculated as $\mathrm{P}=\mathrm{Br}+\mathrm{Bg}-\mathrm{Bm}+\mathrm{Y}$ where Y is catch biomass.

## D. Short-Term Projection

1 ) This section does not explain why the MFDP projection was not carried out from 2005 to 2008 and why this analysis is back in the assessment. It seems that the information available is of sufficient quality to allow this type of projection.
2 ) The projection methodology is appropriate, but stochastic projection, incorporating uncertainty in abundance at age estimates and recruitment estimates would help to evaluate alternative management actions by providing probability of achieving management objectives or risk of exceeding limits.
E. Medium-Term Projections - The text states that Fmsy is provided in Table 7, but it is not (nor can it be from a simple yield-per-recruit analysis).

## F. Long-Term Projections and G. Biological Reference Points

1) The reference points have not been revised. However, considering the changes in the assessment methodology and some evidence Blim should be revised upwards. Some work to investigate a possible change of Blim should be encouraged (or an explanation is required to explain why these reference points should be kept as they are).
2 ) Long-term, stochastic projection should be used to determine the fishing mortality rate associated with $\mathrm{B}_{\lim }$ (as a candidate for $\mathrm{F}_{\mathrm{lim}}$ ) and its uncertainty (to derive $\mathrm{F}_{\mathrm{pa}}$ and potentially a revised $\mathrm{B}_{\mathrm{pa}}$ ) as well as MSY reference points.

## H. Management and ICES Advice

1) The ICA model suggests a recent increase in spawning biomass (2008 SSB $=55800 \mathrm{t}$ ) to greater than Bpa ( 44000 t ), such that a rebuilding program is no longer necessary. However, important caveats should be communicated in the management advice. The ICA calibration is based on a short survey series, and the calibration relationships are weak, with some year effects. The resulting estimates of terminal SSB are imprecise, and much of the uncertainty in terminal SSB is not included in estimates of precision. For example, a large portion of the spawning stock is composed of 1 ringers, for which proportion mature is poorly understood, and geometric mean abundance is assumed in the terminal year.
2 ) Comparison of fishery yields and TACs indicates that the management system can effectively control the fishery (e.g., TAC was slightly exceeded in two years in the last 20 years).
3 ) Previous ICES advice that 'catches of around 5000 t would be associated with stock recovery' appears to be unsubstantiated, because catches have not been that low in the observed catch series.
4 ) 'ACFM Advice' should be updated to 'ICES Advice.'

# Evaluation of proposed rebuilding plan for Celtic Sea herring 

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## Introduction

The herring to the south of Ireland in the Celtic Sea and in Division VIIj comprise both autumn and winter spawning components. For the purpose of stock assessment and management, these areas have been combined since 1982. The stock experienced a collapse in the 1970s, and again, in the mid 2000s.

The ICES advice for 2007, 2008 and 2009 has been that there should be no targeted fishing without a rebuilding plan. In 2008, the local Irish management committee presented a rebuilding plan to the European Commission and Council. The plan was not formally adopted, but the TAC for 2009 was consistent with the plan. Subsequently, in early 2009, the plan was endorsed by the Commission.

The plan (cited below) incorporated scientific advice with the main elements of the EU policy statement. A schematic representation of the plan is shown in Figure 1.

1 ) For 2009, the TAC shall be reduced by $25 \%$ relative to the current year (2008).

2 ) In 2010 and subsequent years, the TAC shall be set equal to a fishing mortality of $\mathrm{F}_{0.1}$.
3 ) If, in the opinion of ICES and STECF, the catch should be reduced to the lowest possible level, the TAC for the following year will be reduced by 25\%.

4 ) Division VIIaS will be closed to herring fishing for 2009, 2010 and 2011.
5 ) A small-scale sentinel fishery will be permitted in the closed area, Division VIIaS. This fishery shall be confined to vessels, of no more than 65 feet in length. A maximum catch limitation of $8 \%$ of the Irish quota shall be exclusively allocated to this sentinel fishery.
6 ) Every three years from the date of entry into force of this Regulation, the Commission shall request ICES and STECF to evaluate the progress of this rebuilding plan.
7) When the SSB is deemed to have recovered to a size equal to or greater than $B_{p a}$ in three consecutive years, the rebuilding plan will be superseded by a long-term management plan.

In March 2009, the European Commission asked ICES (text of Commission interpretation below) to evaluate the plan. ICES was asked to evaluate if points 2 and 3 of the plan were precautionary:

- For 2010 and subsequent years the TAC will be set consistent with a fishing mortality rate of $\mathrm{F}_{0.1}=0.19$.
- If, in the opinion of ICES and STECF the catch should be reduced to the lowest possible level, the TAC for the following year will be reduced by $25 \%$.

This document presents a draft response to the questions posed by the Commission, on the proposed plan. Therefore this document is only concerned with points 2 and 3 of the proposed plan.

## Materials and methods

Evaluation was performed using HCS-Celtic (Skagen, 2009). The program HCS-Celtic (Harvest Control rule Simulation-Celtic Sea herring) is a program for stochastic simulation of harvest control rules. The program is intended to imitate the normal advisory process where the stock is assessed one year before the TAC year. Because of that, a projection is made through the intermediate year to obtain the stock abundance at the start of the TAC year. HCS mimics that process without running actual assessments as part of the simulations. Instead, observation errors are specified as distributions and carried forward in predictions to get the numbers that are the basis for management decisions. Options for implementation error and bias are also available.

The program consists of a population model that generates yearly stock numbers at age, an observation model that transfers the stock numbers into noisy, 'observed' numbers, a decision rule through which a TAC is derived according to the observed stock (projected forward if relevant) and an implementation model that translates the TAC into actual removals. These removals are then input to the population model for the next time step. The outline is shown in Figure 2.

The program is run as a bootstrap, with the following stochastic elements:

- Initial numbers
- Recruitments
- Observation noise
- Implementation noise

This model was an adaptation of the original HCS model (Skagen, 2008). HCS -Celtic included an extra feature to test the effect of zero catch on SSB. If SSB < Blim, then a reduction of $25 \%$ applies. If not, the HCR applies as stated (Skagen, 2009). A subsequent modification was made which derives an SSB for input to the harvest control rule assuming a TAC in the fishing year corresponding to the target F (program available at http://groupnet.ices.dk). This modification showed slightly more conservative results. At low Fs this is not apparent as the HCR is not triggered due to the low exploitation on the stock. At higher Fs the modified version has slightly lower associated risks. Overall both versions produced very similar results. Since the pro-
portion of F prior to spawning is only 0.2 the minor differences between the model versions is understandable. Further runs were carried using the modified version.

The program was used to screen over target F and Btrigger levels. In addition, ranges of \% TAC reductions if SSB < Blim were also explored. Sensitivity analysis was conducted to evaluate observation and implementation error and bias.

## Input Data

The HCS model used the following input data:

- Start Year
- Age range in the population
- Age range for the calculation of average fishing mortality
- For each age: population numbers, natural mortality, selection, catch weight, stock weight and proportion mature.
- TAC in the starting year.
- Parameters for the stock-recruit function.

For these simulations the start year was 2009 as this is the last year that an assessment was carried out. The age range used in the population model was 1-6, with mean fishing mortality calculated over ages 2-5. The 2009 population numbers were from the final assessment in 2009. Following the procedure of the assessment and forecast, 1 ringers were replaced with geometric mean recruitment from 1995-2006. Population numbers of 2 ringers in the intermediate season (2009) were calculated by the degradation of the geometric mean recruitment (1995-2006). Selection at age was taken from the final assessment run in 2009. Natural mortality was assumed to be constant every year. The mean weights in the catch and in the stock are calculated as averages over the last three years (2005-2008). The maturity ogive for this stock assumes that $50 \%$ of 1 ringers, and $100 \%$ of subsequent ringers are mature.

Three estimates of intermediate catch in 2009 were used. This was necessary because the catch in the intermediate year $(2009 / 2010)$ includes the first quarter of the advice (TAC) year. Therefore the TAC set for 2010 influences the intermediate year catch. The interim year catches estimates were as follows:

6,809 15\% increase based on EU TAC Decision Rule for stocks where SSB is increasing (ICES, 2009).
$7,507 \quad 56 \%$ increase, based on $\mathrm{F}_{0.1}(2009)=0.17$.
7,763 71 \% increase, based on $\mathrm{F}_{0.1}(2007)=0.19$.
Apart from the two alternative interim year catches, all of the inputs described above were used in the forecasts that were carried out at the 2009 working group (Table 1).

Table 1. Input data used in the simulations.

|  | Weight in the <br> stock $(\mathrm{kg})$ | Weight in the <br> catch $(\mathrm{kg})$ | Proportion <br> Mature | F | Population <br> Numbers 2009 | Natural <br> Mortality |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.078 | 0.086 | 0.5 | 0.008 | 360168 | 1 |
| 2 | 0.107 | 0.110 | 1 | 0.080 | 131499 | 0.3 |
| 3 | 0.126 | 0.131 | 1 | 0.137 | 145357 | 0.2 |
| 4 | 0.148 | 0.149 | 1 | 0.147 | 44017 | 0.1 |
| 5 | 0.157 | 0.164 | 1 | 0.137 | 84238 | 0.1 |
| 6 | 0.166 | 0.175 | 1 | 0.137 | 22269 | 0.1 |

## Stock Recruit

In all simulations, the same stock-recruit relationship was used. The pattern in the stock-recruit pairs indicated that none of the classical models were appropriate. The segmented regression model was chosen, and applied to data for 1958-2006. Data from the most recent two years were excluded because they are less well estimated. Model fitting was conducted using R (http://www.r-project.org/) using Julios' algorithm (Julios, 2001), see Figure 3. The SSB changepoint was estimated at 41,000 t, and the plateau level of recruitment at 416 million individuals (Table 2). A log-normal distribution of the recruitments was assumed. The distribution was truncated at 0.1 and 3.0 to avoid drawing recruitments far outside the historical range. The modeled and expected distributions of recruitments are shown in the Figure 4. At cumulative probability 0.2 to 0.5 the model predicts higher recruitment than observed, though elsewhere there was excellent agreement. Diagnostics of the model fit are presented in Figure 5.

Table 2. Parameters of the segmented regression model fit for Celtic Sea herring, 1958-2006.

| Slope | Mean recruitment (thousands) | Change point SSB (tonnes) | SSQ | p | S.E |
| :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |
| 10.17 | 416424 | 40,944 | 16.37 | 0.06 | 0.60 |

### 1.1 Results

A number of scenarios were tested, using several runs of the modified HCS-Celtic program. Initial runs investigated a broad range of target F levels (0.2-1.0), trigger biomass ( $B_{\text {trigger }} 26,000-4,000$ ), and \% reductions ( $25 \%-75 \%$ ) when $S S B<B_{\text {trigger. }}$ Contour plots were used to present the results of these simulations, showing target F on the horizontal, and $B_{\text {trigger }}$ on the vertical, with separate rows for the different $\%$ reductions, and columns for year combinations. The legend of probabilities ( $\mathrm{SSB}<\mathrm{Blim}_{\mathrm{lim}}$ ) is provided on the right, in terms of colours. According to ICES common practice, levels of less than $5 \%$ are considered to be in accordance with the precautionary approach.

Subsequent runs (Table 3) tested a narrower range of target Fs , in the range of $\mathrm{F}_{0.1}$. $\mathrm{F}_{0.1}$ has been estimated as 0.17 (ICES, 2009) and 0.19 (ICES, 2007). Therefore target F in the range 0.17 to 0.19 was evaluated. These runs were to simulate the rebuilding plan as is requested by the Commission. These runs also considered the sensitivities of various factors.

## Initial runs

Results of initial runs are shown in Figures 6.1 (without bias), 6.2 (with bias). Summary plots are shown in Figure 6.3. The results of initial screening (Run 1) showed that target Fs above 0.4 have increased risk of $\mathrm{SSB}<\mathrm{B}_{\text {trigger }}$. These simulations suggest that target $F$ in the range up to 0.4 is precautionary at any chosen trigger biomass to $45,000 \mathrm{t}$ and any $\%$ reduction from $25 \%$ to $75 \%$. Unacceptable risks are associated with target $F$ above 0.6 . At high target $F$, risk is lower when higher $B_{\text {trigger }}$ is chosen. The inclusion of implementation bias ( $10 \%$ ) did not alter the risk profile appreciably (Figure 6.2). From Figure 6.3 it can be seen that risk to Blim is predicted to increase to unacceptable levels by 2012 and that target F in the range 0.6 to 1.0 is predicted to lead to $\mathrm{B}_{\mathrm{lim}}$ at some point in the simulation period. Figure 6.4 shows risk profiles for the first and second 10-year periods of the simulation. There was increased risk associated with higher target F , and lower $\mathrm{B}_{\text {trigger }}$ in the second period.

Initial screening suggested that target F , in the range of recently proposed $\mathrm{F}_{0.1}$ estimates, is precautionary. Subsequent simulations concentrated on a range of F in this region. The base case scenario that was tested considered three $\mathrm{F}_{0.1}$ estimates, $\mathrm{B}_{\text {trigger }}=$ Blim from the proposed rebuilding plan and the proposed percentage reduction when SSB < Blim (Run 3). A 10\% implementation bias was considered appropriate and CV on the observation and implementation models was fixed at $20 \%$. This was based on an interim year catch of $7,763 \mathrm{t}(\mathrm{F}=0.19)$.

Figure 7.1 shows trajectories of realised F and yield, SSB and risk to Blim for this run, and Appendix 1 contains detailed outputs. This F range is associated with minimal risk ( $<1 \%$ ) to Blim, and a building of the stock to levels where yields of about $13,000 \mathrm{t}$ are realised over the latter part of the simulation period. The highest target F (ICES, 2007; $\mathrm{F}_{0.1}$ ) does not increase the risk to any appreciable degree and is associated with similar yields.

Table 3. Details of simulation runs conducted.

|  |  | Int |  |  | \% TAC <br> Run <br> redn. | Notes |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Type |  |  |  |  |  |
| 1 | Broadscale | 7763 | $0.2-1.0$ | $24-44 \mathrm{~K}$ | $25-75$ |  |
| 2 | Broadscale | 7763 | $0.2-1.0$ | $24-44 \mathrm{~K}$ | $25-75$ |  |
| 3 | Base case | 7763 | $0.17-0.19$ | $26-44 \mathrm{~K}$ | 25 | IAV 5-25\% |
| 4 | Sensitivity | 7763 | $0.17-0.19$ | $26-44 \mathrm{~K}$ | 25 | Obs CV $0.2-0.4$ |
| 5 | Sensitivity | 7763 | $0.17-0.19$ | $26-44 \mathrm{~K}$ | 25 | Obs. Bias -0.1 to 0.5 |
| 6 | Sensitivity | 7763 | $0.17-0.19$ | $26-44 \mathrm{~K}$ | 25 | Imp. CV 0.1 to 0.3 |
| 7 | Sensitivity | 7763 | $0.17-0.19$ | $26-44 \mathrm{~K}$ | 25 | Imp. bias 0.1 to 0.3 |
| 8 | Sensitivity | 7763 | $0.17-0.19$ | $26-44 \mathrm{~K}$ | 25 |  |
| 9 | Sensitivity | 7507 | $0.17-0.19$ | $26-44 \mathrm{~K}$ | $25-75$ |  |
| 10 | Sensitivity | 6809 | $0.17-0.19$ | $26-44 \mathrm{~K}$ | $25-75$ |  |

The sensitivity of the base case run was tested against several factors, namely:

- Inter-annual TAC variation (Runs 4 )
- Precision and bias on observation model (Runs 5 and 6 respectively)
- Precision and bias on implementation model (Runs 7 and 8 respectively)
- Interim year catch (Runs 9 and 10)

It was not appropriate to investigate the effect of changing the \% reduction of TAC when SSB < Blim. This was because the range of $\mathrm{F}^{\prime}$ s in the region of $\mathrm{F}_{0.1}$, which is specified in the proposed plan do not bring the SSB below Blim or the other trigger points chosen.

The risk profiles for these sensitivity runs are presented in Figures 8.1 to 8.3. In the $F$ and $B_{\text {trigger }}$ region of the proposed plan, no IAV was associated with a risk to Blim, that was appreciably lower than the base case (Figure 8.1). The proposed rebuilding plan appears robust to a plausible range of implementation errors and biases. Slightly higher risk was found to be associated with an observation CV of $40 \%$ (Figure 8.2). None of the likely interim year catches alter the risk profile to any extent (Figure 8.3).

## Discussion

## Results of simulations for conformity with the precautionary approach

The simulations conducted in this exercise predict that the proposed rebuilding plan is consistent with the precautionary approach to fisheries management. Target F in the range of recent estimates of $\mathrm{F}_{0.1}$ is not associated with risk of $\mathrm{SSB}<\mathrm{B}_{\lim }$.

The $25 \%$ TAC reduction when $\mathrm{SSB}<\mathrm{B}_{\lim }$ was shown to be precautionary, when target $\mathrm{F}<0.4$. At higher target F , acceptable risks were associated with a $75 \% \mathrm{TAC}$ reduction, and Btrigger in range of $40,000 \mathrm{t}$ to $45,000 \mathrm{t}$. These simulations are based on the best estimate of current (2009) stock size, and low historic catch levels. However, if the stock was decreasing and catches were at levels observed historically, then it is not clear if a $25 \%$ reduction would be precautionary.

Point 3 may not be appropriate for a long term management plan for this stock. The current simulations are only relevant to the proposed rebuilding plan. However following this plan there will be minimal risk to Blim. Thus, the overall rule in the plan is precautionary, if the target $F$ in point 2 is followed.

## Interpretation of the rebuilding plan

The clause "the catch should be reduced to the lowest possible level, the TAC for the following year will be reduced by $25 \%$ ", in point 3 of the plan may lead to confusion. This is a standard wording used in the EU policy statements on the fixing of catch opportunities. The stakeholder committee intended this text to represent the scenario, when the scientific advice is for a zero-catch. This would apply if the SSBTAC year $\leq$ Blim. In this scenario, the plan provides for a $25 \%$ reduction in TAC, not TAC $=0$ (Figure 3).

## Stock dynamics and the population model

It is important that the stock dynamics are well understood and an adequate basis for simulating the plan. The underlying population model was that of the 2009 accepted ICES assessment (ICES, 2009b). This was considered an extension of the 2007 benchmark assessment of this stock. The independent reviewers endorsed the decisions
made and concluded that the assessment was methodology was "generally sound", but with "some inconsistencies" (Cadrin et al. 2009). These inconsistencies were as follows:
a. Short survey series, with weak relationships to canum and some year effects.
b. A large portion of the spawning stock composed of 1-ringers, that are poorly estimated.
c. Assumption of constant selectivity.
d. Assumed selectivity at the oldest ages for the entire time series may lead to misinterpretation of the apparent shifts in age selectivity by the fishery.
e. No estimates of discards.
f. Consumption of herring as forage not estimated.
g. Mixed-stock resource and connectivity with adjacent management units assessment should be developed.

Points $\mathrm{a}, \mathrm{b}, \mathrm{c}$, and d above represent structural aspects of the fisheries data that cannot be improved on at this time. The survey data used represents the longest time series of comparable surveys available. The poor estimation of 1-ringers can only be improved when a recruit series is available. The population model assumes $50 \%$ maturity at 1 -ringer. This is a compromise. It is known that more than $50 \%$ of fish in Celtic Sea catches are mature at 1-ring (Lynch, 2009). However these are probably fast growing recruits. On the other hand, slower growing fish, present in the Irish Sea (Brophy and Danilowicz, 2002), may have a later maturity. The selection assumption seems valid for the separable period assumed in the model as the fishery pattern has been relatively constant over this period.

This evaluation is comparable with others conducted recently on West of Scotland herring, NEA mackerel and western horse mackerel, where discarding was either not accounted for at all, or else not fully accounted for in the observation model.

The consumption of herring as forage by predatory marine animals has not been evaluated in the population model. No good estimates of herring consumption exist in this area. Natural mortality in many herring stocks is poorly understood. Recent agreed management plans for west of Scotland herring and horse mackerel were based on data that did not explicitly consider the forage consumption. Though the approach taken for Celtic Sea herring is broadly comparable with other stock assessments, it is clear that more work needs to be done on the level of consumption of herring as forage.

Point f of the review group's comments is considered to add uncertainty to the stock dynamics. The current stock assessment model does not consider, the effect the mixing of juveniles, and indeed adults, of this stock with the neighbouring Irish Sea stock. One approach to this problem would be to employ an assessment model such as Roel et al. (2009). Another approach would be to use the framework developed by Kell et al. (2009). It was intended to use this framework to evaluate the proposed rebuilding plan. However, insufficient time was available to develop the program, which is currently not fully operational.

## Stock recruitment relationship and recruitment variability

The stock recruitment data do not suggest that any of the classical models (Beverton and Holt, Ricker, Shepherd) are applicable. The data show low and high recruitments
at low and high stock size. The segmented regression was chosen, and provided an independent estimate of the changepoint SSB, below which recruitment impairment is considered to occur. The estimate of change point $(41,000 \mathrm{t})$ is close to recent estimates (45,000 t; STECF, 2006; 47,000 t; Clarke and Egan, 2008). Recent recruitment has fluctuated around a mean level of 360 million, lower than the long term mean estimated by the segmented regression model ( 416 million). The stock recruitment relationship may produce higher recruitments in the simulation period, than have been observed in the recent past. However, sensitivity analysis suggests that results are robust error and bias in the observation model.

## Progress towards Fmsy by 2015

According to the political commitment at the World Summit on Sustainable Development at Johannesburg, in September 2002, fish stocks should be maintained at or restored to levels that can produce maximum sustainable yield, not later than 2015. The current exercise did not seek to estimate Fmsy. However, F0.1 can be used as a proxy for $\mathrm{F}_{\text {ms }}$.

## Conclusions

In answer to the specific questions posed by the Commission the following answers can be given:

1 Setting a TAC, consistent with a fishing mortality rate of $\mathrm{F}_{0.1}=0.19$, for 2010 and subsequent years is not associated with an unacceptable risk of SSB < Blim, in the simulation period 2009-2029.

2 If TACs consistent with F in the range 0.17 to 0.19 are set, then there is minimal risk that $\mathrm{SSB}<\mathrm{B}_{\lim }$ in the simulation period 2009-2029. However at fishing takes place at $\mathrm{F}>0.4$ the $25 \%$ TAC reduction in the proposed plan may not be precautionary.

The proposed rebuilding plan for Celtic Sea and Division VIIj herring is estimated to be in accordance with the precautionary approach, if the target fishing mortality of $\mathrm{F}_{0.1}$ is adhered to.

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Figure 1. Schematic representation of the decision rule in the proposed rebuilding plan. $B_{\text {triger }}=$ $B_{\text {lim }}=26,000 \mathrm{t}$.


Figure 2. Outline of the simulation loop in the HCS-Celtic program.

## Stock Recruit Relationship



Figure 3. Stock recruit relationship using segmented regression and SSB changepoint indicated.

Cumulated Distribution of Recruitment


Figure 4. Stock recruit relationship cumulative distribution of observed and expected recruitments.


Figure 5. Stock recruit relationship diagnostics of the model fit.


Figure 6.1. Contour plot showing the probability that $\operatorname{SSB}<B_{l i m}$. Run 1, broad scale screening without implementation bias. The $x$-axis shows potential target $F$ over a broad range, and the $y$ axis the differing levels of trigger biomass. Each line represents a \% TAC reduction, to be implemented if SSB $<\mathrm{B}_{\lim }(\mathbf{2 5 \%}, 50 \%$ and $75 \%)$.


Figure 6.2. Contour plot showing the probability that $\mathrm{SSB}<\mathrm{B}_{\mathrm{lim}}$. Run 2, broad scale screening with implementation bias. The $x$-axis shows potential target $F$ over a broad range, and the $y$-axis the differing levels of trigger biomass. Each line represents a \% TAC reduction, to be implemented if SSB $<\mathrm{B}_{\lim }(\mathbf{2 5 \%}, \mathbf{5 0 \%}$ and $\mathbf{7 5 \%}$ ).


Figure 6.3. Trajectory plots for broad scale screening exercise (Run 2). Simulated trajectories for F, SSB, catch and risk to $B_{\text {lim. }}$ Precautionary 5 \% risk level indicated.


HCS_Celtic Sea Herring : Prob(SSB<Blim) 2019-2029


Figure 6.4. Plot showing the probability that SSB < Blim (Run 2). Risk of being below Blim in the first and second 10 -year periods of the simulation. The $x$-axis shows potential target $F$ over a broad range, and the $y$-axis the differing levels of trigger biomass. Each line represents a \% TAC reduction, to be implemented if SSB <Blim $(\mathbf{2 5 \%}, 50 \%$ and $75 \%)$.


Figure 7.1. Trajectory plots for simulations of proposed rebuilding plan (Run 3). Simulated trajectories for F, SSB, catch and risk to $B_{\text {lim. }}$. Risk to $B_{\text {lim }}$ presented as a histogram because of very low levels.


Figure 8.1. Plot showing the probability that $\operatorname{SSB}<\operatorname{Blim}_{\text {lim }}$ for sensitivity analysis of the base case to differing inter-annual TAC variations. The $x$-axis shows potential target $F$, and the $y$-axis the differing levels of trigger biomass. Each line represents a \% TAC reduction, to be implemented if SSB < Blim. $_{\text {lin }}$.


Figure 8.2. Plot showing the probability that $\operatorname{SSB}<\operatorname{Bim}_{\mathrm{lim}}$ for sensitivity analysis of the base case to observation and implementation model error (CV) and bias. The x-axis shows potential target $F$, and the $y$-axis the differing levels of trigger biomass. Top left to bottom right: Runs 8 (implementation bias), 6 (observation bias), 7 (implementation CV) and 5 (observation CV).



Figure 8.3. Plot showing the probability that SSB < Blim for sensitivity analysis of the base case to the three most likely interim year catches. The $x$-axis shows potential target $F$, and the $y$-axis the differing levels of trigger biomass. Each plot represents a separate interim year catch.

Appendix 1. Detailed output of base case simulation run.

| Target F | Btrig | Year | F | SSB | Catch | TAC | Change | Fiim | Pquash |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.17 | 26000 | 2009 | 0.159 | 61502 | 7763 | 7763 | 0 | 0 | 0 |
|  |  | 2010 | 0.193 | 61689 | 9423 | 9244 | 8.6 | 0.004 | 0 |
|  |  | 2011 | 0.195 | 70392 | 11287 | 11065 | 17.2 | 0 | 0 |
|  |  | 2012 | 0.19 | 68904 | 10693 | 10692 | -2.8 | 0 | 0 |
|  |  | 2013 | 0.198 | 72535 | 11926 | 11606 | 7.1 | 0.001 | 0 |
|  |  | 2014 | 0.195 | 73619 | 11920 | 11743 | 2.8 | 0.001 | 0 |
|  |  | 2015 | 0.191 | 74682 | 11866 | 1182 | 0.7 | 0 | 0 |
|  |  | 2016 | 0.195 | 75930 | 12249 | 12054 | 1.6 | 0 | 0 |
|  |  | 2017 | 0.194 | 76897 | 12597 | 12246 | 2.5 | 0 | 0 |
|  |  | 2018 | 0.191 | 77019 | 12442 | 12241 | -0.6 | 0 | 0 |
|  |  | 2019 | 0.192 | 77461 | 12632 | 12471 | 2.1 | 0 | 0 |
|  |  | 2020 | 0.198 | 77431 | 13026 | 12698 | 2 | 0 | 0 |
|  |  | 2021 | 0.195 | 77316 | 12759 | 12382 | -2.1 | 0 | 0 |
|  |  | 2022 | 0.196 | 77193 | 12674 | 12355 | 0.1 | 0 | 0 |
|  |  | 2023 | 0.192 | 77547 | 12679 | 12467 | 0.4 | 0 | 0 |
|  |  | 2024 | 0.194 | 77436 | 12719 | 12534 | 0.4 | 0 | 0 |
|  |  | 2025 | 0.192 | 77832 | 12775 | 12508 | -0.2 | 0 | 0 |
|  |  | 2026 | 0.193 | 78015 | 12686 | 12444 | 0.5 | 0 | 0 |
|  |  | 2027 | 0.195 | 79414 | 12977 | 12815 | 2.9 | 0 | 0 |
|  |  | 2028 | 0.191 | 78531 | 12635 | 12417 | -3.2 | 0 | 0 |
|  |  | 2029 | 0.19 | 78626 | 12700 | 12459 | 0.2 | 0 | 0 |
| 617 | 35000 | 2009 | 0.158 | 62161 | 7763 | 7763 | 0 | 0 | 0 |
|  |  | 2010 | 0.198 | 62176 | 9653 | 9558 | 10.9 | 0 | 0 |
|  |  | 2011 | 0.195 | 71179 | 11336 | 11200 | 15.6 | 0.001 | 0 |
|  |  | 2012 | 0.193 | 6\%54 | 10999 | 10824 | -2.4 | 0 | 0 |
|  |  | 2013 | 0.185 | 73415 | 11304 | 11276 | 3.9 | 0 | 0 |
|  |  | 2014 | 0.195 | 74954 | 12227 | 11906 | 6 | 0 | 0 |
|  |  | 2015 | 0.2 | 75970 | 12604 | 12331 | 3.4 | 0.001 | 0 |
|  |  | 2016 | 0201 | 75865 | 12541 | 12211 | -0.6 | 0.001 | 0 |
|  |  | 2017 | 0.197 | 76350 | 12633 | 12337 | 1.1 | 0.001 | 0 |
|  |  | 2018 | 0.194 | 76575 | 12568 | 12237 | -0.3 | 0.001 | 0 |
|  |  | 2019 | 0.198 | 76705 | 12705 | 12449 | 1.2 | 0.001 | 0 |
|  |  | 2020 | 0.191 | 77267 | 12354 | 12029 | -2.3 | 0 | 0 |
|  |  | 2021 | 0.195 | 79062 | 12773 | 12519 | 2.7 | 0.001 | 0 |
|  |  | 2022 | 0.192 | 79410 | 12691 | 12391 | 0 | 0.001 | 0 |
|  |  | 2023 | 0.195 | 78514 | 12910 | 12778 | 2.9 | 0.001 | 0 |
|  |  | 2024 | 0.198 | 79436 | 13131 | 12679 | -0.9 | 0.001 | 0 |
|  |  | 2025 | 0.194 | 78079 | 12804 | 12485 | -1.2 | 0 | 0 |
|  |  | 2026 | 0.192 | 78053 | 12675 | 12344 | -0.7 | 0 | 0 |
|  |  | 2027 | 0.196 | 78357 | 12976 | 12697 | 2.3 | 0 | 0 |
|  |  | 2028 | 0.2 | 77815 | 13160 | 12850 | 0.9 | 0 | 0 |
|  |  | 2029 | 0.196 | 77217 | 12940 | 12524 | -1.6 | 0 | 0 |
| 0.17 | 44000 | 2009 | 0.16 | 61166 | 7763 | 7763 | 0 | 0 | 0 |
|  |  | 2010 | 0206 | 61309 | 9806 | 9554 | 12.8 | 0.001 | 0 |
|  |  | 2011 | 0.201 | 69457 | 11266 | 11098 | 13.2 | 0 | 0 |
|  |  | 2012 | 0201 | 68078 | 10891 | 10801 | -1.6 | 0.003 | 0 |
|  |  | 2013 | 0.201 | 71328 | 11722 | 11402 | 4.3 | 0.002 | 0 |
|  |  | 2014 | 0.197 | 72793 | 11764 | 11597 | 3 | 0.002 | 0 |
|  |  | 2015 | 0.208 | 78896 | 12520 | 12141 | 4.3 | 0.002 | 0 |
|  |  | 2016 | 0.199 | 74811 | 1282 | 12049 | -1 | 0.003 | 0 |
|  |  | 2017 | 0.197 | 75515 | 12380 | 12139 | 1.3 | 0.002 | 0 |
|  |  | 2018 | 0.199 | 76039 | 12588 | 12367 | 1.3 | 0.002 | 0 |
|  |  | 2019 | 0.204 | 76230 | 12801 | 12266 | -0.8 | 0.005 | 0 |
|  |  | 2020 | 0.2 | 76422 | 12606 | 12488 | 0.8 | 0.003 | 0 |
|  |  | 2021 | 0.199 | 76731 | 12673 | 12288 | -0.9 | 0.003 | 0 |
|  |  | 2022 | 0.196 | 76987 | 12444 | 12175 | -0.2 | 0.002 | 0 |
|  |  | 2023 | 0.197 | 77622 | 12714 | 12572 | 2.3 | 0.002 | 0 |
|  |  | 2024 | 0.197 | 77580 | 12828 | 12577 | 0.4 | 0.002 | 0 |
|  |  | 2025 | 0.2 | 77501 | 12865 | 12616 | -0.1 | 0.001 | 0 |
|  |  | 2026 | 0.196 | 77411 | 12800 | 12478 | -1 | 0.001 | 0 |
|  |  | 2027 | 0.195 | 77423 | 12626 | 12405 | -0.2 | 0 | 0 |
|  |  | 2028 | 0.196 | 77582 | 12729 | 12566 | 0.9 | 0.002 | 0 |
|  |  | 2029 | 0.192 | 77572 | 12504 | 12313 | -1.5 | 0.001 | 0 |

Appendix 1. (continued).

| Target | Btrig | Year | F | SSB | Catch | TAC | Change | Fim | Parash |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.18 | 26000 | 2009 | 0.158 | 61961 | 7763 | 7763 | 0 | 0.001 | 0 |
|  |  | 2010 | 0.207 | 61747 | 9984 | 9774 | 139 | 0.002 | 0 |
|  |  | 2011 | 0.207 | 69857 | 11805 | 11651 | 16.4 | 0.001 | 0 |
|  |  | 2012 | 0.215 | 67963 | 11609 | 11299 | -28 | 0.001 | 0 |
|  |  | 2013 | 0.209 | 70715 | 12062 | 11822 | 4.9 | 0.001 | 0 |
|  |  | 2014 | 0.207 | 71432 | 12243 | 11995 | 2.2 | 0 | 0 |
|  |  | 2015 | 0.205 | 72214 | 12300 | 12015 | 0.3 | 0 | 0 |
|  |  | 2016 | 0.203 | 72940 | 12265 | 12157 | 2.2 | 0.001 | 0 |
|  |  | 2017 | 0.204 | 73863 | 12432 | 12357 | 1.8 | 0.001 | 0 |
|  |  | 2018 | 0.212 | 74594 | 13111 | 12682 | 1.5 | 0.001 | 0 |
|  |  | 2019 | 0.204 | 74773 | 12634 | 12438 | -08 | 0 | 0 |
|  |  | 2020 | 021 | 75049 | 13081 | 12809 | 1.9 | 0 | 0 |
|  |  | 2021 | 0.206 | 74991 | 12930 | 12739 | 0 | 0 | 0 |
|  |  | 2022 | 0.206 | 74894 | 12903 | 12531 | -1 | 0 | 0 |
|  |  | 2023 | 021 | 75197 | 13047 | 12636 | 1.2 | 0 | 0 |
|  |  | 2024 | 0.2 | 75018 | 12537 | 12392 | -2.4 | 0 | 0 |
|  |  | 2025 | 0.207 | 75175 | 13022 | 12715 | 2.1 | 0 | 0 |
|  |  | 2026 | 0.213 | 75124 | 13279 | 12719 | 0.2 | 0 | 0 |
|  |  | 2027 | 0.202 | 75163 | 12733 | 12565 | -02 | 0.001 | 0 |
|  |  | 2028 | 0.205 | 75291 | 12882 | 12664 | 0.6 | 0 | 0 |
|  |  | 2029 | 021 | 75682 | 13223 | 12908 | 1.5 | 0.001 | 0 |
| 0.18 | 35000 | 2009 | 0.158 | 62319 | 7763 | 7763 | 0 | 0 | 0 |
|  |  | 2010 | 0.209 | 62201 | 10182 | 9941 | 159 | 0.001 | 0 |
|  |  | 2011 | 0.209 | 70065 | 11948 | 11742 | 139 | 0 | 0 |
|  |  | 2012 | 0.212 | 68181 | 11628 | 11438 | -09 | 0.001 | 0 |
|  |  | 2013 | 0.211 | 70939 | 12111 | 11868 | 3.6 | 0.001 | 0 |
|  |  | 2014 | 0.219 | 71455 | 12706 | 12380 | 5.2 | 0.003 | 0 |
|  |  | 2015 | 0.214 | 72318 | 12603 | 12364 | 0.2 | 0.002 | 0 |
|  |  | 2016 | 0.205 | 72791 | 12242 | 12042 | -19 | 0.003 | 0 |
|  |  | 2017 | 0.208 | 73357 | 12700 | 12351 | 2.2 | 0.002 | 0 |
|  |  | 2018 | 0.209 | 73584 | 12887 | 12406 | 0.3 | 0.001 | 0 |
|  |  | 2019 | 0.204 | 73590 | 12522 | 12298 | -07 | 0.001 | 0 |
|  |  | 2020 | 021 | 74294 | 12818 | 12696 | 3 | 0 | 0 |
|  |  | 2021 | 0.213 | 74799 | 13126 | 12777 | 0.8 | 0.001 | 0 |
|  |  | 2022 | 0.205 | 74549 | 12766 | 12606 | -17 | 0.001 | 0 |
|  |  | 2023 | 0.199 | 75385 | 12535 | 12437 | -0.4 | 0 | 0 |
|  |  | 2024 | 0.205 | 76199 | 13009 | 12631 | 2 | 0 | 0 |
|  |  | 2025 | 0.208 | 75732 | 13195 | 12895 | 1.1 | 0 | 0 |
|  |  | 2026 | 0.205 | 75359 | 12931 | 12720 | -0.4 | 0 | 0 |
|  |  | 2027 | 0.209 | 75084 | 13205 | 12842 | 0 | 0 | 0 |
|  |  | 2028 | 0.204 | 74756 | 12792 | 12566 | -18 | 0 | 0 |
|  |  | 2029 | 0.208 | 74884 | 12984 | 12709 | 1.4 | 0 | 0 |
| 0.18 | 44000 | 2009 | 0.158 | 62136 | 7763 | 7763 | 0 | 0.001 | 0 |
|  |  | 2010 | 0.206 | 62044 | 10058 | 9908 | 158 | 0.001 | 0 |
|  |  | 2011 | 0.209 | 70258 | 11871 | 11712 | 158 | 0 | 0 |
|  |  | 2012 | 0.216 | 68625 | 11841 | 11532 | -09 | 0.002 | 0 |
|  |  | 2013 | 0.212 | 71257 | 12209 | 12037 | 4 | 0.003 | 0 |
|  |  | 2014 | 0.216 | 71702 | 12517 | 12186 | 1.6 | 0.004 | 0 |
|  |  | 2015 | 0.217 | 72282 | 12601 | 12248 | 0.8 | 0.006 | 0 |
|  |  | 2016 | 0.216 | 72974 | 12772 | 12330 | 0 | 0.006 | 0 |
|  |  | 2017 | 0.219 | 73408 | 12857 | 12567 | 1.8 | 0.006 | 0 |
|  |  | 2018 | 0.214 | 73806 | 12814 | 12604 | 0.3 | 0.005 | 0 |
|  |  | 2019 | 0.212 | 74001 | 12793 | 12677 | 0.5 | 0.005 | 0 |
|  |  | 2020 | 0.216 | 74563 | 12954 | 12656 | 0 | 0.005 | 0 |
|  |  | 2021 | 0.213 | 74946 | 13016 | 12748 | 0.4 | 0.007 | 0 |
|  |  | 2022 | 0.213 | 74829 | 12973 | 12706 | -0.7 | 0.006 | 0 |
|  |  | 2023 | 0.216 | 74868 | 13177 | 12773 | 0.7 | 0.005 | 0 |
|  |  | 2024 | 0.212 | 74675 | 12878 | 12689 | -03 | 0.005 | 0 |
|  |  | 2025 | 0.208 | 74882 | 12907 | 12681 | -1 | 0.005 | 0 |
|  |  | 2026 | 0.214 | 74741 | 13116 | 12736 | 0.6 | 0.006 | 0 |
|  |  | 2027 | 0.211 | 74515 | 13038 | 12857 | 0.4 | 0.003 | 0 |
|  |  | 2028 | 0.214 | 74642 | 13190 | 12868 | 0 | 0.002 | 0 |
|  |  | 2029 | 0.212 | 74829 | 12912 | 12710 | -0.5 | 0.001 | 0 |

## Appendix 1.(continued).

| Target F | Btrig | Year | F | SSB | Catch | TAC | Change | Plim | Pcrash |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.19 | 26000 | 2009 | 0.155 | 62930 | 7763 | 7763 | 0 | 0 | 0 |
|  |  | 2010 | 0.218 | 62427 | 10797 | 10621 | 21.8 | 0.002 | 0 |
|  |  | 2011 | 0.22 | 69566 | 12448 | 12284 | 13.3 | 0 | 0 |
|  |  | 2012 | 0.214 | 66959 | 11571 | 11491 | -5 | 0.003 | 0 |
|  |  | 2013 | 0.216 | 69767 | 12143 | 11928 | 3.1 | 0.002 | 0 |
|  |  | 2014 | 0.22 | 70570 | 12618 | 12418 | 5.7 | 0.002 | 0 |
|  |  | 2015 | 0.22 | 71204 | 12904 | 12485 | -1.1 | 0.002 | 0 |
|  |  | 2016 | 0.211 | 71328 | 12410 | 12252 | -0.3 | 0.001 | 0 |
|  |  | 2017 | 0.218 | 72048 | 12934 | 12674 | 3.1 | 0.001 | 0 |
|  |  | 2018 | 0.216 | 72394 | 12927 | 12645 | -0.2 | 0 | 0 |
|  |  | 2019 | 0.217 | 72584 | 13012 | 12723 | 1.1 | 0 | 0 |
|  |  | 2020 | 0.215 | 72824 | 12943 | 12788 | 0.4 | 0 | 0 |
|  |  | 2021 | 0.22 | 72858 | 13222 | 12991 | 1.4 | 0.001 | 0 |
|  |  | 2022 | 0.216 | 72998 | 13034 | 12861 | -0.6 | 0.001 | 0 |
|  |  | 2023 | 0.214 | 73454 | 12978 | 12754 | -1 | 0.002 | 0 |
|  |  | 2024 | 0.218 | 74060 | 13268 | 12946 | 1.8 | 0.001 | 0 |
|  |  | 2025 | 0.223 | 73937 | 13500 | 13164 | 1.5 | 0.001 | 0 |
|  |  | 2026 | 0.22 | 73794 | 13416 | 13108 | -0.1 | 0.001 | 0 |
|  |  | 2027 | 0.219 | 73582 | 13280 | 13027 | -0.6 | 0 | 0 |
|  |  | 2028 | 0.221 | 73266 | 13362 | 13066 | -0.5 | 0.001 | 0 |
|  |  | 2029 | 0.219 | 73091 | 13167 | 12907 | -0.3 | 0.001 | 0 |
| 0.19 | 35000 | 2009 | 0.161 | 61051 | 7763 | 7763 | 0 | 0.001 | 0 |
|  |  | 2010 | 0.22 | 61015 | 10540 | 10379 | 19 | 0.003 | 0 |
|  |  | 2011 | 0.233 | 68491 | 12621 | 12308 | 16.6 | 0.001 | 0 |
|  |  | 2012 | 0.229 | 66282 | 11850 | 11594 | -4.5 | 0.004 | 0 |
|  |  | 2013 | 0.228 | 69327 | 12488 | 12140 | 3.5 | 0.007 | 0 |
|  |  | 2014 | 0.224 | 70431 | 12614 | 12424 | 2.8 | 0.006 | 0 |
|  |  | 2015 | 0.227 | 70732 | 12916 | 12602 | 1.4 | 0.007 | 0 |
|  |  | 2016 | 0.229 | 70821 | 13123 | 12767 | 1.3 | 0.005 | 0 |
|  |  | 2017 | 0.232 | 70922 | 13238 | 12829 | 0.5 | 0.006 | 0 |
|  |  | 2018 | 0.224 | 70706 | 12806 | 12624 | -1 | 0.005 | 0 |
|  |  | 2019 | 0.218 | 70840 | 12635 | 12419 | -1.3 | 0.004 | 0 |
|  |  | 2020 | 0.223 | 71334 | 12959 | 12793 | 2.4 | 0.002 | 0 |
|  |  | 2021 | 0.22 | 71533 | 12854 | 12604 | -0.6 | 0.001 | 0 |
|  |  | 2022 | 0.222 | 71750 | 12919 | 12698 | 0.4 | 0.002 | 0 |
|  |  | 2023 | 0.223 | 71954 | 13141 | 12742 | 1.1 | 0.004 | 0 |
|  |  | 2024 | 0.224 | 72042 | 13040 | 12767 | 0.1 | 0.002 | 0 |
|  |  | 2025 | 0.219 | 72501 | 12811 | 12712 | 0 | 0.002 | 0 |
|  |  | 2026 | 0.22 | 72875 | 13143 | 12883 | 0.9 | 0.002 | 0 |
|  |  | 2027 | 0.217 | 72736 | 13081 | 12898 | 0 | 0.001 | 0 |
|  |  | 2028 | 0.218 | 72837 | 13124 | 13065 | 1.3 | 0.001 | 0 |
|  |  | 2029 | 0.216 | 73021 | 12975 | 12823 | -1.3 | 0.002 | 0 |
| 0.19 | 44000 | 2009 | 0.157 | 62139 | 7763 | 7763 | 0 | 0.001 | 0 |
|  |  | 2010 | 0.222 | 61604 | 10720 | 10541 | 20.3 | 0.002 | 0 |
|  |  | 2011 | 0.224 | 68915 | 12341 | 12128 | 14.2 | 0.002 | 0 |
|  |  | 2012 | 0.236 | 66397 | 12333 | 12154 | 0.3 | 0.001 | 0 |
|  |  | 2013 | 0.237 | 68375 | 12455 | 12227 | 0.1 | 0.003 | 0 |
|  |  | 2014 | 0.234 | 68522 | 12481 | 12149 | 0.8 | 0.006 | 0 |
|  |  | 2015 | 0.231 | 69238 | 12661 | 12516 | 2.5 | 0.005 | 0 |
|  |  | 2016 | 0.226 | 69949 | 12465 | 12256 | -1.4 | 0.004 | 0 |
|  |  | 2017 | 0.222 | 70267 | 12612 | 12483 | 1.1 | 0.005 | 0 |
|  |  | 2018 | 0.226 | 70671 | 12927 | 12757 | 1.9 | 0.004 | 0 |
|  |  | 2019 | 0.225 | 71082 | 12910 | 12547 | -1 | 0.003 | 0 |
|  |  | 2020 | 0.224 | 71498 | 12898 | 12757 | 1.4 | 0.003 | 0 |
|  |  | 2021 | 0.226 | 71389 | 13023 | 12789 | -0.6 | 0.002 | 0 |
|  |  | 2022 | 0.229 | 72163 | 13101 | 12746 | 0 | 0.002 | 0 |
|  |  | 2023 | 0.235 | 72323 | 13562 | 13120 | 1.6 | 0.004 | 0 |
|  |  | 2024 | 0.231 | 72388 | 13239 | 13095 | 0.2 | 0.003 | 0 |
|  |  | 2025 | 0.233 | 72393 | 13390 | 13114 | 0.3 | 0.004 | 0 |
|  |  | 2026 | 0.234 | 72414 | 13564 | 13198 | 0.7 | 0.005 | 0 |
|  |  | 2027 | 0.228 | 72125 | 13235 | 13001 | -2 | 0.005 | 0 |
|  |  | 2028 | 0.233 | 72359 | 13399 | 13028 | 1 | 0.005 | 0 |
|  |  | 2029 | 0.226 | 72192 | 12965 | 12737 | -1.8 | 0.003 | 0 |

## Review of Celtic Sea management plan

John Simmonds 9th June 2009

## The proposed management plan for Celtic Sea (Zones VIIhjk):

1 ) For 2010 and subsequent years the TAC will be set consistent with a fishing mortality rate of $\mathrm{F} 0.1=0.19$.
2 ) If, in the opinion of ICES and STECF the catch should be reduced to the lowest possible level, the TAC for the following year will be reduced by $25 \%$
(With additional restrictions not tested)

## The EC requested the evaluation should address:

1) the consequences of implementing the above rule instead of implementing ICES' current advice for this stock according to the precautionary approach;
2 ) the extent to which the application of this rule would deliver management inconformity with the precautionary approach;
3 ) the extent to which the application of this rule would deliver maximum sustainable yield from the stock;
4 ) where possible, stochastic future time-streams of TACs and fishing effort
5 ) necessary to catch those TACs should be made available to STECF for economic analysis.

## ICES has requested a review based on the 8 clauses

1 ) Is the study based on a correct interpretation of the management plan*?
2 ) Have the authors presented the correct information for evaluating the precautionary nature of the plan?
3 ) Are the assumed stock dynamics an adequate basis for simulating the plan?
4 ) Are the assumed fleet dynamics an adequate basis for predicting future catches and fishing mortality in the simulation?
5 ) Has an appropriate model formulation been used?
6 ) Have all sources of process and estimation error that could impact the conclusions been adequately represented?
7 ) Are the authors' conclusions valid?
8 ) Has the request been answered in full?
The review below is organised around these eight clauses and a brief look at the questions raised by the EC.

## 1. Correct interpretation

The plan appears to be correctly interpreted, though the diagram in Figure 1 is a poor representation of the process.

## 2. Correct information on the performance of the plan.

Figure 7.1 provides the basic information on the precautionary performance of the plan as simulated. Blim is specified at $26,000 \mathrm{t}$ on the basis of lowest observed SSB, however, with a well established breakpoint at $41,600 \mathrm{t}$ in the $\mathrm{S} / \mathrm{R}$ relationship (see below) there are indications this may be miss specified. Nevertheless current specification is at 26,000 and ICES criteria of 5\% are correctly dealt with.

## 3. Adequacy of the assumed stock dynamics

The fitted $\mathrm{S} / \mathrm{R}$ relationship and input data are given in Figures 3 and 5. A example of the simulated data are given in Figure 4. The fitted model appears to be an adequate single model description of the historic stock and recruitment. The choice of model is plausible and well supported by the data, the fitting method is suitable to give a good fit between data and model. The SSB breakpoint is well described by the data and the value of SSB at the breakpoint and mean recruitment above the breakpoint are well established. The diagnostics indicate that the $S / R$ data are stable over time and therefore the use of the fitted model to infer the future is reasonable. There are indications of slight deviation from the model below the breakpoint, but this does not substantively influence any aspect of the results.

The clipping of simulated values appears appropriate (Fig 4) but its unclear over which biomass values the comparison of simulated and predicted have been compared (normal practice would be to use only the observed SSBs hopefully this is what is presented). The report states that the model predicts higher recruitment from 0.2 to 0.5 , but Fig 4 seems to show the reverse. No mention is made of a year on year correlation in recruitment, though this seems evident in the timeseries.

In conclusion the $S / R$ model appears well founded with the exception of autocorrelation and thus may be classed as marginally adequate.

## 4. Adequacy of assumed fleet dynamics for predicting future catches and fishing mortality in the simulation?

No description is provided to describe fishery dynamics, though the use of implementation CV and bias of $20 \%$ and $10 \%$ seem reasonable (or too uncertain) given the recent history of fishing. The mismatch between TAC year and fishery and assessment implies some flexibility between years. Evaluations of between year flexibility (Methods 2008) suggest this is not a problem for low exploitation rates such as those proposed. Control through TAC would seem to be effective based on data in the ICES stock summary sheet (ICES 2008)

## 5. Appropriateness of model formulation

The software used was supplied from the ICES website and has been validated by use on other similar simulations.

Parameterisation of the model is rather superficial, ignoring any autocorrelation in either recruitment or measurement error, though both are available in the software. This simple approach does not include an evaluation of assessment error, or correlation in that error. Given that the evaluation was for one specific harvest rule with only one survey to tune the assessment and that assessment model used (FLICA) is available in FLR it would be feasible to carry out a fuller evaluation. At
the least it would be helpful to check the characteristics of error in the assessment in a small number of runs of a single case.

## 6. Have all sources of process and estimation error that could impact the conclusions been adequately represented?

The basis for fishery dynamics and the implementation error is poorly described in the report. This base case assumes $20 \%$ CV on implementation and observation, and a $10 \%$ bias on implementation. While these figures are plausible, there is little presented in the document to back this up. Examination of the ICES ACOM advice sheet for Celtic Sea herring indicates that the catches recorded (and included in the data used for the $\mathrm{S} / \mathrm{R}$ model) are below the TAC in the last 12 years. Suggesting that the values assumed for bias and CV on implementation may overestimate these errors.

The recent survey seems to perform well but the timeseries may be a bit short to determine errors well. Once the survey is compared with a converged VPA errors may be more reliably established, though this will not be possible for several more years. In this case choice of $20 \%$ CV may be over optimistic for a single vessel acoustic survey.
A sensitivity analysis to observation and implementation bias is provided, these vary from from -0.1 to 0.5 and .1 to 0.3 respectively. Similarly observation and implementation CV is varied from 0.2 to 0.4 and 0.1 to 0.3 respectively. These provide a fairly simple and effective approach to evaluating if the results are critical to the assumptions.

This sensitivity analysis shows risks are not significant except for observation CVs of 0.4 at Fs above 0.185. However, the investigation does not cover higher CVs and higher implementation bias combined.
Nevertheless the insensitivity of the conclusions to plausible if simplistic errors is such that this is an adequate approach in these circumstances. This would not be the case if the results were more marginal.

## 6. Validity of authors' conclusions

The authors conclude that from the current starting point fishing at F0.1 ( 0.17 to 0.19 ) is in accordance with the precautionary approach, this is supported by the analysis.

Some concern is expressed that previously observed catches may be too high to sustain a $25 \%$ restriction on TAC. However, if the target F of $\sim 0.18$ a $25 \%$ is complied with (within $10 \%$ bias and $20 \% \mathrm{CV}$ ) then the $25 \%$ should be acceptable unless a very long run of poor recruitment occurs. However, because autocorrelation has not been included in the simulated $\mathrm{S} / \mathrm{R}$ relationship and SSB does not fall below Blim during simulations at $\mathrm{F}=0.17-0.19$ the consequences of applying $25 \%$ with SSB below Blim are not tested and not know.

The conclusions are based on the current PA points which might benefit from reevaluation.

## 7. Has the request been answered in full?

In addition to the questions raised by ICES the EC asked the following:

1. the consequences of implementing the above rule instead of implementing ICES' current advice for this stock according to the precautionary approach;

The rebuilding plan proposed might result in slower recovery than that obtainable by following ICES precautionary advice of no directed fishery. However, currently this ICES advice does not seem to be being followed (see ICES stock summary sheets), so the plan may be lead to recovery more quickly than NOT following ICES precautionary advice.
2. the extent to which the application of this rule would deliver management in conformity with the precautionary approach;

The simulations show that the plan is precautionary within the ICES definition (risk $<5 \%$ SSB below Blim). The evaluation was limited in scope ignoring some aspects that may be important but as a sensitivity analysis using more demanding conditions of bias and error was carried out and also show acceptable performance. Thus given the relatively low exploitation rate $(\mathrm{F}=\sim 0.18)$ limitations are acceptable and the conclusion that the plan is precautionary is reasonable.

The inclusion of a $25 \%$ inter-annul restriction on reduction in TAC given SSB below Blim is not testable within the range of recruitment simulated, as the F target does not bring the stock to Blim to allow this to be tested. Any test if this would be artificial. Nevertheless it is expected that such a restriction is acceptable as the exploitation rate implied by F-0.18 is lower than the $25 \%$ restriction thus TACs should come down faster than the stock.
3. the extent to which the application of this rule would deliver maximum sustainable yield from the stock;

The plan is designed to give fishing at F0.1. Based on yield per recruit studies presented in HAWG (ICES 2009), Celtic Sea herring has no defined Fmax within a plausible range of F. In the absence of Fmax, F0.1 forms a good surrogate for Fmsy.

# Review of the Celtic Sea Herring management plan evaluation 

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## 1. Is the study based on a correct interpretation of the management plan*?

This study integrates mainly points 2 and 3 of the EU policy statements.
Point 2: "In 2010 and subsequent years, the TAC shall be set equal to a fishing mortality of F0.1." $\mathrm{F} 0.1=0.17$ and 0.19 were integrated as input parameters for the simulations therefore this point is fully implemented into the modeling framework used for this study.

Point 3: The discussion section of this study clearly explains how point 3 "the catch should be reduced to the lowest possible level, the TAC for the following year will be reduced by $25 \%$ " has been interpreted in the study. The wording of the EU policy statement is misleading but the interpretation of this clause appears to be correct: if SSBtacyear<Blim , A $25 \%$ reduction in TAC is applied.

Points 4 and 5 deal with the closure of VIIaS. Considering the possible migrations between areas, it would have been interesting to make an attempt to consider how the closure of this area may or may not reflect on the short term scale on VIIj. That, however, could have added another factor of uncertainties.

Points 1, 6 and 7 were not points to be discussed/implemented in the management plan.
2. Have the authors presented the correct information for evaluating the precautionary nature of the plan?

The authors focused on the following criteria to evaluate the plan : the probability of SSB falling below Blim, Realized F, catch, SSB and risk to Blim. Those parameters are sufficient to understand what the plan may imply for the stock and its harvesting.

The presentation of the results however lacks of a sound temporal limit which also reflects the lack of temporal limit in implementation of the EU policy statements. From a modeler point of view, on figure 6.3, most runs tend towards some steady states situation past 2019 and until the end of the runs in 2029. The occurrence of steady-state situation in nature is itself a nearly philosophical debate among scientists. Here, considering the variation in recruitment on a short term species and all the potential factors that may affect the fishing efforts (adjustment in fleets due to gas price for example), estimating the state of the stock without too much errors is probably only possible for a few number generations (e.g. 3-4). In that sense, going past 2019 may be misleading as a quick reading of the plots may suggest people outside of the modeling world that things will be nicely steady after 2019 which is not realistic as this situation is more a signal from the model structure rather than the natural variations of its parameters (biology, harvesting, environmental factors...). This could be avoided by limiting all plots to 2019.

## 3. Are the assumed stock dynamics an adequate basis for simulating the plan?

The discussion section of this study mentions the problems identified in the assessment methodology (i.e. benchmark). The general impression is that the authors are aware of the problems and of some of the solutions to explore to solve them. However, no exploratory/sensitivity analyses were made to evaluate the impact of those factors. The reason was mainly the lack of knowledge/data of some aspects or the fact some of those points were usually not considered into the assessments. The stock dynamics does not include the effect of mixing from adjacent area as this point is considered to add some uncertainty. However, this discussion gives the overall impression that the stock dynamics is well relatively described considering the current knowledge on the fisheries.

Dealing with the stock recruitment relationship is a source of issues due to the variability of the recruitment between years. The authors mentioned any of the classical models were applicable but no indication of the quality of the fitting of those relationships was provided (such as $\mathrm{r}^{2}$ or AIC). A similar S-R situation has been observed with the Bay of Biscay anchovy long term management plan (STECF, 2008) and none of the regular models were apparently applicable as well. In practice, all models for BoB anchovy had the same AIC which meant that none of the model was performing better than the others therefore the adjustment was quite poor. In that case, choosing the most convenient (i.e the least worst) model may be based on some criteria such as the shape or the number of parameters. Here, the segmented regression is simpler to use than any curvilinear approach and still provides more or less the same (poor) adjustment.

A closer look at figure $3(\mathrm{~S}-\mathrm{R})$ shows the stock has been quite low in recent years but recruitment has been very variable. Data go back to 1958 and up to 2006. Considering the recent biomass of the stock, fitting the S-R relationships for the whole time series may put too much weight on past biomass situations. The authors explain however that the fitting on the whole time series provides some robust results.

Under the scrutiny of different neutral eyes, any attempt to fit a relationship on those points may result in different approaches/point of view. It would have been interesting to have a better explanation of this choice of segmented regression and maybe a sensitivity analysis using another descriptor/relationship for S-R.

## 4. Are the assumed fleet dynamics an adequate basis for predicting future catches and fishing mortality in the simulation?

The fleet dynamics is not explicitely mentioned in this study as the total catch is ruled by the Harvest Control Rules (HCR) through the automatic TAC set by the estimates of SSB. The activity of the fleet is not simulated. The closure of VIIaS and its possible effects on the redeployment of the local fleet are not integrated into this study as well. Considering the distances between areas, it is hard to tell if integrating that measure into this plan would have affected the biomass in VIIj.

## 5. Has an appropriate model formulation been used?

The general procedure used for this study is more or less standard and is described with details in the "Materials and Method" section. From an outsider point of view,
the information is detailed enough to understand the approach taken. The procedure itself seems fine for the task (the same principles have been applied on some other MSE e.g. North Sea Cod).

## 6. Have all sources of process and estimation error that could impact the conclusions been adequately represented?

The major sources of uncertainties for this stock have been reviewed in this study in the discussion section (see answers \#3 and \#4).

One source of error apparently not accounted into this study remains using the most recent assessment as a starting point for the simulations. The Benchmark report (Cadrin et al., 2009) mentions that the assessment model lacks of performance diagnostics (e.g. the screening of possible retrospective pattern). Considering the various sources of uncertainties from the data and from the performance of the model, it would have been interesting to test the behavior of the Management plan for different starting years (for example, the assessments for the last 5 years). I suspect as the management plan converge towards steady-state on the long run that the plan, overall, is not strongly affected by some possible biases in the initial conditions of the simulation. However, as this plan is more oriented towards a short-term situation, there may be some significant changes in the first years simulated.

## 7. Are the authors' conclusions valid?

The modeling framework provides the usual responses one can expect from looking for the compromise between protecting and harvesting the stock: setting higher Btrigger is a protective measure and lowering F reduce the risk of collapse. The management plan is qualitatively sound on these mandatory aspects. Overall, the author's conclusions from the simulations appear to be valid. The sensitivity analysis which goes through various value of F and reductions of TAC is helpful to test how precautionary the $25 \%$ TAC reduction rule and F target (0.17-0.19) are in comparison to other values.

In the discussion, the paragraph "point 3 may not be..." is too concise to explain why the $25 \%$ TAC reduction should not be in a long term management plan.

Note: on figures 8.1, 8.2, 8.3, it would have been helpful to tell that probabilities are very low. For the quick reader, it seems like no plot has been correctly produced.

## 8. Has the request been answered in full?

The request has been answered in full for points 2 and 3 which were the only points ICES was asked to evaluate. The authors dealt with the consequence for catch and stock biomass of the implementation of the points 2 and 3 , the limit of use of those rules to remain within the precautionary approach and to sustain maximum yield. The simulations also provide time series of the evolution of SSB, F, TAC and risk to be below Blim (although as I wrote earlier, the temporal extension after 2019 might be misleading). All those points were in the EU request letter*.

Through the sensitivity analysis, the authors went further than the initial request by exploring the risks associated with a wide range of target $\mathrm{F}(0.2-1)$ and higher reduc-
tion of TAC ( $50-75 \%$ ). This kind of initiative is also within the EU request letter as ICES was "invited to propose alternative rules or modified rules on its own initiative or in consultation with RACs and to evaluate these"
*:(see
http://groupnet.ices.dk/HAWG2009/Celtic\ Sea\ Herring/Commision\ reply\  Celtic\%20Sea\%20 RebPlan.pdf )

## References:

Anon., 2008. Report of the STECF Meeting on long-term anchovy management. STECF. 77p + annex.

Cadrin, S., Pawlowski, L., Goethel, D. and Kerr, L. 2009. Benchmark review of Celtic Sea Herring. Unpublished report to ICES ACOM. 9 pp.

## Annex 13 Technical Minutes of the Celtic Sea Review Group (RGCS)

## 2009

26 May - 4 June 2009, Fairhaven Massachusetts, USA

Reviewers: Steve Cadrin (chair), Adam Barkley, Greg DeCelles, Dan Goethel, Nikki Jacobson, Lisa Kerr, Dave Martins, Cate O'Keefe, Sally Roman, Tony Wood
Working Groups:

- Working Group on Celtic Seas Ecoregion (WGCSE, Colm Lordan chair)
- Herring Assessment Working Group (HAWG, Maurice Clarke chair)
- Working Group on the Assessment of Hake Monk and Megrim (WGHMM, Carmen Fernandez, chair)

Secretariat: Barbara Schoute

Process - The ICES advisory service quality assurance program requested that a team of graduate and post-doctoral students and their professor serve as a review group. The group initially met on 26 May to review the ICES advisory process, RG guidelines and to assign several WG report sections to each reviewer. A second meeting was held on 27 May to review standard ICES assessment models (XSA, ICA, SURBA, TSA and BADAPT). Members reviewed WG report sections independently, then presented their summaries and reviews to the group in a series of meetings during 13 June to discuss reviewers' proposals and form RG conclusions.

General - Stock assessment reports for 32 stocks were reviewed (Table1). The WG reports were generally informative, and WG decisions about data, model choice and specification and interpretations were clearly explained and justified. The RG concludes that the reports are technically correct, and the RG agrees with WG recommendations, with few exceptions. In nearly all cases, the assessments appropriately applied the procedures specified in the stock annexes. Some general issues were raised for many stocks related to discards, definition of assessment and management units and standardized methods. These general observations should be considered for the next benchmark reviews of these stocks.

Table 1. Stocks reviewed ordered by working group (WG), terms of reference (ToR), type of assessment and assessment method.

| WG | Stock | Name | ToR | type | method |
| :---: | :---: | :---: | :---: | :---: | :---: |
| wgcse | cod-7e-k | Cod in Divisions VIIe-k (Celtic Sea Cod) | Update | no method | Benchmarked |
| wgcse | cod-iris | Cod in Division VIla (Irish Sea) | Update | assess | BADAPT |
| wgcse | cod-rock | Cod in Division VIb (Rockall) | No assessment | no advice |  |
| wgcse | cod-scow | Cod in Division VIa (West of Scotland) | Update | assess trends | TSA |
| wgcse | had-7b-k | Haddock in Divisions VIIb-k | Update | assess trends | XSA |
| wgcse | had-iris | Haddock in Division VIla (Irish Sea) | Update | assess trends | SURBA |
| wgcse | had-rock | Haddock in Division Vlb (Rockall) | Update | assess | XSA |
| wgcse | had-scow | Haddock in Division Vla (West of Scotland) | Update | assess | TSA |
| wgcse | whg-7e-k | Whiting in Divisions VIIe-k | Same Advice | assess trends | XSA |
| wgcse | whg-iris | Whiting in Division VIla (Irish Sea) | Same Advice | assess trends | SURBA |
| wgcse | whg-scow | Whiting in Division VIa (West of Scotland) | Update | assess | SURBA |
| wgcse | ple-7h-k | Plaice in Divisions VIIh-k (Southwest of Ireland) | Same Advice | catch trends | - |
| wgcse | ple-celt | Plaice in Divisions VIIf,g (Celtic Sea) | Update | assess | XSA |
| wgcse | ple-echw | Plaice in Division VIIe (Western Channel) | Update | catch trends | XSA |
| wgcse | ple-iris | Plaice in Division VIIa (Irish Sea) | Update | assess | ICA |
| wgcse | sol-celt | Sole in Divisions VIIf, g (Celtic Sea) | Update | assess | XSA |
| wgcse | sol-echw | Sole in Division VIIe (Western Channel) | Update | survey trends | Benchmarked |
| wgcse | sol-iris | Sole in Division VIIa (Irish Sea) | Update | assess | XSA |
| wgcse | nep-11 | Nephrops in Division VIa (North Minch, FU 11) | Update | assess trends | Benchmarked |
| wgcse | nep-12 | Nephrops in Division Vla (South Minch, FU 12) | Update | assess trends | Benchmarked |
| wgcse | nep-13 | Nephrops in Division Vla (Firth of Clyde, FU 13) | Update | assess trends | Benchmarked |
| wgcse | nep-14 | Nephrops in Division VIIa (Irish Sea East, FU 14) | No assessment | assess trends |  |
| wgcse | nep-15 | Nephrops in Division VIIa (Irish Sea West, FU 15) | Update | assess trends | Benchmarked |
| wgcse | nep-17 | Nephrops in Division VIIb (Aran Grounds, FU 17) | Update | assess trends | Benchmarked |
| wgcse | nep-19 | Nephrops in Division VIIa, g, (South East \& West of IRL, FU 19) | No assessment | assess trends |  |
| wgcse | nep-2022 | Nephrops in Division VIIf,g,h (Celtic Sea, FU 20-22) | No assessment | assess trends |  |
| wgcse | nep-7bcj | Nephrops in Division VIIb,c,j,k (Porcupine Bank, FU 16) | No assessment | assess | Status changed |
| wgcse | ang-ivvi | Anglerfish in Division Ila, IIIa, Subarea IV and VI | Update | assess trends | - |
| wgcse | meg-scrk | Megrim in Subarea VI (West of Scotland and Rockall) | Update | catch trends | - |
| wghmm | ang-78ab | Anglerfish in Divisions VIIb-k and VIIIa,b,d | Update | assess trends | - |
| wghmm | mgw-78 | Megrim in Divisions VIIb-k and VIIIa,b,d | Update | survey \& cpue trends | - |
| hawg | her-irls | Herring in Division VIIa South VIIg,h,j,k (Celtic Sea \& S. Ireland) | Benchmark | assess trends | ICA |
| hawg | her-irlw | Herring in Divisions VIa (South) and VIIb,c | Same Advice | assess trends | ICA |
| hawg | her-nirs | Herring in Division VIIa North of $52^{\circ} 30^{\prime} \mathrm{N}$ (Irish Sea) | Same Advice | assess trends | - |
| hawg | her-vian | Herring in Division VIa (North) | Update | assess | ICA |

Most of the stocks that were reviewed are caught in mixed-stock fisheries. Many assessments include mixed-stock considerations, estimate discards, and include them in the stock assessment. However, the treatment of discards varies widely among assessments. The RG recommends that all information on discarded catch should be reported, the magnitude of discards should be estimated or approximated for all fleets, and if the proportion of discards is substantial, discards should be included as a component of catch for the entire assessment series for exploratory analyses and possibly as the basis for fishery management advice. The RG recognizes that estimates of discards for some fleets and in historical periods will be highly uncertain. However, many of the stocks in this group have substantial discards, and retrospective patterns suggest under-reported catch. The RG concludes that including discard approximations may improve the accuracy and consistency of assessments.

The definition of assessment units and management units do not correspond for many stocks in this group. Many management areas include multiple assessment units, such that catch of each assessment unit is not directly managed, because TACs can be taken from any component stock. Assessment and management unit definitions should be re-evaluated to improve the effectiveness of management. Furthermore, stock units should reflect biological stocks within the practical constraints of fishery monitoring and resource surveys for stocks that overlap. Many of the datapoor assessments in this group may benefit from aggregation of management units

## Stock: Herring in Divisions VIIa South VIIg,h,j,k (Celtic Sea)

(see May 2009 benchmark review, Appendix A)
Assessment Type: Benchmark
Assessment: Analytical
Forecast:
short-term projections
Assessment method: ICA
Consistency the assessment is relatively consistent, but stock status has changed.
Stock Status: SSB recently increased ( 2008 SSB=55 800t) to greater than Bpa (44000t), such that a rebuilding program is no longer necessary. However, there is considerable uncertainty in the stock assessment.

Management Plan: There is no explicit management plan
General and Technical Comments: see Appendix A

## Conclusions:

The ICA calibration is based on a short survey series, and the calibration relationships are weak, with some year effects. The resulting estimates of terminal SSB are imprecise, and much of the uncertainty in terminal SSB is not included in estimates of precision. For example, a large portion of the spawning stock is composed of 1-ringers, for which proportion mature is poorly understood, and geometric mean abundance is assumed in the terminal year.

## Stock: Herring in Divisions VIIa (South) and VIIb,c

| Assessment Type: | Same advice as last year. |
| :--- | :--- |
| Assessment: | Trends |
| Forecast: | None presented |

Assessment method: No final assessment has been accepted for this stock. A separable VPA without a tuning index was used in an exploratory assessment of this stock (terminal F scenarios $=0.2,0.4,0.5,0.6$ ). An ICA with acoustic survey tuning index was used in 2006 and 2007, but subsequently discontinued.

Consistency: The stock status is considered to be the same as last year. Retrospective assessments were conducted for each terminal F scenario, terminal F of 0.4 to 0.5 produced more stable estimates of SSB and F.

Stock Status: The current SSB is uncertain, but thought to be below Blim. Current F is uncertain, but thought to be above Flim (Flim = 0.33). Estimates of SSB from an acoustic survey (in 2008 the survey took place in July designed to survey summer feeding aggregations) estimated abundance at $43,000 \mathrm{t}$. It is unclear whether the stock is contained in this area during this time of year.

Management Plan: The TAC for this stock in 2009 is $9,314 \mathrm{t}$ (this represents a $20 \%$ decrease from 2008). ICES advises that a rebuilding plan should be put in place to reduce F and catch, however, there is no explicit management plan currently in place for this stock. .

## General Comments:

- Results of the assessment indicate that SSB and recruitment have declined and F has increased since late 1980s.
- Log catch ratios confirm that mortality has been increasing in recent years.
- There are no recruitment indices for this stock; however, abundance of young age classes (1-ringers) in catches by the fishery has been at its lowest in recent years.
- Additionally, age composition of catch indicates an apparent truncation of the age structure in this stock.
- Because of the limited number of years of data collection and changes in the design of the acoustic survey it is not considered to be a useful tuning index for the assessment.
- Herring are caught in a targeted single-species fishery. There is a lack of information on discards for this stock and bycatch in the fishery. Since the fishery is predominantly ( $>90 \%$ of TAC) taken by Ireland, a priority should be put on obtaining discard and bycatch data from the Irish fleet. Additionally, there is a large amount of misreporting of herring catch by area ( $\sim 23$ to $29 \%$ of total catch in recent years).
- Ecosystem information was not considered in examination of stock trends.


## Technical Comments:

- An exploratory assessment was conducted using a separable VPA (Lowestoft VPA software) in which 4 choices for terminal fishing mortalities ( $0.2,0.4,0.5,0.6$ ) were examined. An assessment of retrospective bias indicated that recent $F$ is
most likely between 0.4 and 0.5 . The current SSB is uncertain but most likely less than Bpa and Blim. No tuning indices were used, therefore, data from most recent, unconverged years is not considered informative. Yield per recruit analysis was conducted in 2006 ( $\mathrm{F} 0.1=0.17$ ), it was not considered necessary to update this analysis.
- The current reference points are listed in the annex which states that Bpa is 81,000 t and Blim is 110,000 , these numbers appear to be inverted and should be corrected.
- An effort should be made to ensure tables and figures are easily interpretable. The font should be increased in tables and figures. Figures should be revised to ensure that axes are readable. Figures 6.2.1.1, 6.4.1.1, 6.6.2.3 were unreadable.
- The current assessment does not give a strong basis for issuing advice. A tuning index is needed for this stock assessment to be useful in estimating SSB and F in recent years.


## Conclusions:

The assessment has been performed as indicated in the annex. The lack of a tuning index for this stock makes recent information uninformative and estimates of recent SSB and F are uncertain. All indicators point to SSB being below critical levels (Blim). The RG agrees with the need for a rebuilding plan for this stock. Given the current uncertainty with respect to stock status a more precautionary approach in the management of this stock, specifically a more aggressive reduction in TAC ( $>20 \%$ ), seems appropriate.

## Stock: Herring in Division VIIa North (Irish Sea)]

| Assessment Type: | Same as last year |
| :--- | :--- |
| Assessment: | None |
| Forecast: | None |
| Assessment method: | None |

Consistency: a status quo TAC is proposed and is considered precautionary.
Stock Status: SSB > Bpa. There is no Fpa or Flim.
Management Plan: None

## General Comments:

- No discard data is included, but in Annex 8 it is stated that slippage discarding has been increasing over the time series since 1980. It may be appropriate to develop a method for obtaining better estimates of discards especially if the majority of the catch is coming from two vessels in the third quarter of the fishing season. The reason discards were not included is because of the variability of discarding rates observed with rates ranging from 20 percent in 1982 to 50 percent in 1991.
- The fact that there is mixing between the Celtic Sea and the Irish Sea stocks may lead to inflated biomass estimates for younger age classes from the acoustic survey for the VIIa stock. The otolith work done seems like an appropriate method for determining the abundance of Celtic Sea juveniles in the Irish Sea, but does need to be done on a yearly basis as proposed by the WG. The divergence between the acoustic and catch-at-age data from landings may be diminished when stock separation is taken into account.
- Sampling of the smaller Mourne fishery may want to be included. The fishery's landings have increased over the last couple of years and in the document it is stated that this fishery may be fishing on a subpopulation of the stock.
- Ecosystem considerations were discussed in Annex 8, but not included in the report. The Annex stated that a rise in temperature may affect recruitment.
- No mixed fishery concerns were addressed, but do not need to be considered because the herring fishery is a relatively clean fishery in regards to other species.


## Technical Comments:

- The use of weight-at-age data collected from landings data in the third quarter of each fishing year is appropriate because the method captures the weight of the herring in that fishing year which accounts for the decrease in weight-at-age observed over the time series.
- The TAC partitioned amounts of $3,500 t$ to the UK and $1,250 t$ to the Republic of Ireland do not add up to the total TAC of 4,800 t.
- The first sentence on page 362- "The TAC for VIIa (N) is partitioned as 3500 t to the UK and 1250 t to the Republic of Ireland." is a repeat of previous sentences in section 7.1.1 and seems out of place. It is not in a paragraph.
- In section 7.2.4 it may be appropriate to include the time series average because it is not stated in the table referenced. The CV of .98 could also be included in the paragraph because it is the highest CV in the time series.
- Figure 7.2.2 B is missing the label B in the upper left corner of the bottom image.
- In Figure 7.2.4 there are 28 trawl length class histograms, but in Figure 7.2.2 A and Section 7.2.3 there are 27 trawls completed that accompanied the acoustic survey. Axis labels could be better identified.
- In Figure 7.2.5 there is no legend to indicate what the changes in color mean.


## Conclusions:

The RG agrees with the WG on the lack of an assessment and the precautionary TAC recommendation. The TAC allows for flexibility in biomass levels as juveniles from the Celtic Sea stock leave VIIa. Further work should be done to address the mixing of stocks and juvenile residence issues to allow for the models discussed in Annex 8 to be applied with better results.

## Stock: Herring in Division Vla North (West of Scotland)

Assessment Type: Update

Assessment:

## Forecast:

Analytical
Both short and medium term forecasts were presented in the assessment.

Assessment method: An integrated catch analysis (ICA) was used and calibrated with the herring acoustic survey.

Consistency: ICA settings have not changed since previous assessments, data was updated but with no significant changes, and the perception of the stock remains relatively unchanged. Retrospective analysis indicates no patternsand bias appears relatively low. The assessment appears relatively noisy, yet balanced.

Stock Status: SSB has been above Btrigger (75,000t) for most of the time series and F (.16) remains well below Ftarget (.25).

Management Plan: As of December 2008 the newly accepted management plan is:

| $\mathrm{F}=0.25$ if SSB $>75000 \mathrm{t}$ | $20 \%$ TAC constraint. |
| :--- | :--- |
| $\mathrm{F}=0.20$ if $\mathrm{SSB}<75000 \mathrm{t}$ but $>62500 \mathrm{t}$ | $20 \%$ constraint on TAC change. |
| $\mathrm{F}=0.20$ if SSB $<62500 \mathrm{t}$ but $>50000 \mathrm{t}$ | $25 \%$ constraint on TAC change |
| $\mathrm{F}=0 \quad$ if SSB $<50000 \mathrm{t}$. |  |

In accordance with this plan, the proposed TAC for 2009 is $21,760 t$, which represents a slight increase from the 2008 TAC of $27,200 \mathrm{t}$.

## General Comments:

- Overall, the assessment is very well done and although it is noisy it appears to be relatively unbiased.
- The RG agrees with the WG that the issues of misreported catch and the high variability in abundance estimates from the acoustic survey that is used to tune the model are the main sources of uncertainty that must be addressed to improve assessment results.
- The issue of misreporting appears to be an enforcement problem that cannot be directly resolved from a modeling standpoint. Although catch estimates are adjusted in order to attempt to resolve this problem, it appears almost impossible to accurately judge misreporting rates. It is suggested that better documentation is made of how catch is adjusted and the reasoning behind such adjustments. Although the method maybe sound, little documentation is given, this makes it difficult for outside sources to determine how or why catch levels were adjusted. In addition, it is suggested that sensitivity runs are made for different levels of catch in order to determine how greatly model outputs differ according to varying degrees of misreporting.
- The variability seen in the acoustic survey is very disconcerting. Relatively little information is given about the general survey protocol. More details should be given either in the stock annex or the assessment document regarding how the acoustic survey is carried out and how it is included in the assessment. Work
should be done to attempt to standardize the acoustic survey in order to reduce the variability in biomass estimates. For instance, it is suggested that the same vessel is used each year, that the survey takes place while a majority of fish are in spawning aggregations, and general sampling design should be standardized. It is believed that such improvements will reduce variability and avoid extreme yearly changes in biomass estimates, such as the $165 \%$ increase in biomass from 2007 to 2008. The impact of such variability on model results is severe. Additionally, the RG agrees that the strong trends in the model fit to the survey over the last few years are quite troubling. Most survey residuals since 2004 have been strongly negative, yet in 2008 all residuals become positive. Possible reasons for these trends should be investigated.
- The use of an age-varying M is seen as appropriate due to the extreme differences in mortality seen in most herring species over their lifecycle, but data supporting the M values used should be given in the annex. Also, it is probably inappropriate to use time-invariant values of M as the ecosystem has changed drastically since 1957, both in terms of predator and prey levels, but also in regards to general environment and temperature regime, all of which affect natural mortality levels. It is suggested that a MSVPA model may be appropriate to update M estimates, as was used in Celtic Sea Herring assessments.
- The RG agrees that the biological sampling for this stock is very poor and should be increased immediately. Only 13 samples were taken in 2008, 11 of which were from the same fleet in the same quarter (Scotland, 3rd quarter). Sampling should be increased and diversified so that samples are taken throughout the year from all fleets.
- Although discards are generally very low and thus not included in the model, it appears that high-grading could be an important issue especially in the freezer trawler fleet. This should be investigated and high grading should be treated as discards and included in future assessments if deemed to be high enough to affect model results.
- In general, it seems that the herring fishery is a clean, directed fishery with little bycatch of other species. Also, a number of ecosystem considerations were discussed regarding the importance of herring as one of the key foundations of the marine food chain. As suggested earlier, it maybe worthwhile to look at MSVPA type models of herring in order to address their importance to the food web and how changes in the ecosystem can affect herring mortality.


## Technical Comments:

Overall, this is an excellent document. Graphs and tables are well labeled and documented. In addition, they all have good descriptions of what is presented. The formatting is also well done with graphs being large enough to read and tables not being too crowded. The variety and types of diagnostic plots is an outstanding feature. The only criticism is that color coding may improve the readability of some graphs, especially the retrospective plots.

## Conclusions:

- The RG agrees with the WG that the ICA assessment is an acceptable model for Herring west of Scotland in area VIa North. Although the assessment is noisy
due to variability in the acoustic survey and relatively high levels of misreporting of catch, it appears to be relatively unbiased. It is suggested that future work be done to increase biological sampling and standardize the acoustic survey, which should help to reduce the noise in the model.
- The RG disagrees with the management plan and the suggestion that the TAC remain at status quo levels. The RG proposes that the precautionary reference points are not in fact precautionary. Bpa is set at 75,000t, which the stock has been above for its entire time series except for a handful of one year intervals in the early 1990s and a period in the late 1970s. At the same time, though, stock abundance has decreased compared to historical estimates and recruitment is at the lowest levels ever seen, while F remains at intermediate levels. It appears that productivity of this stock has declined, possibly due to environmental changes, which will hinder its ability to recover from high fishing levels. It is suggested that target reference points should be reevaluated, and that target SSB should be increased.
- As a foundation of the ecosystem it is important to avoid a stock collapse of herring and so it seems that a more precautionary approach should be undertaken. Also, the ecosystem relations between species and between herring recruitment and the environment should be investigated. In addition, stocks in neighboring areas are in poor conditions and since large proportions of catch in VIa north are from these areas it would make sense to decrease the VIa north TAC to avoid stock collapses in all areas.
- Although stock structure was recently re-evaluated by the WESTHER project, there could be high connectedness between adjacent populations. Mixing between populations could be one reason for the noise in the acoustic survey. It might be better to model these areas as one population with one TAC or to develop a movement model that models each unit as a separate population, but allows movement between areas. This would require data on movement such as a tagging study. Finally, it is suggested that modeling fleets within VIa north separately might be appropriate since each uses different techniques and fishes different geographic locations, and thus will have very different selectivity patterns. This would most likely be best addressed through a forward projecting statistical catch at age model.


# Appendix A. Benchmark Review of Celtic Sea Herring 

6 May 2009 by correspondence

Reviewers:
Steve Cadrin, USA (chair)
Lionel Pawlowski, France
...with assistance from Daniel Goethel and Lisa Kerr, USA

## Summary

Methodology of the Celtic Sea Herring stock assessment was reviewed according to the stock annex (Annex 05 Celtic Sea and Division VIIj_V1_09.doc), and the most recent application of the methodology was reviewed according to the 2009 stock assessment (04-Celtic Sea and Div. VIIj_Herring_2009.doc). Our review addresses the following questions from the ICES Secretariat (B. Schoute 22 April 2009):

1. Is the new assessment methodology correct, of high standard and does it make optimal use of the available data?
2. Are the settings of the forecast (only short term, medium and long term are not relevant here) chosen correctly and do the reference points still apply.
3. Is the methodology adequately described in the stock annex, meaning that the assessment can in principle be carried out by experienced outsiders on the basis of this text.
1) We conclude that the new assessment methodology is generally sound, but some inconsistencies between surveys remain, results are somewhat uncertain, and assumptions of the model should be further explored. The assessment results are relatively consistent, but calibration relationships are weak and based on a short survey time series. Among the alternative assessment models explored, the revised configuration of Integrated Catch Analysis (FLICA) appears to be most appropriate model of data from the fishery and resource. However, we encourage the assessment Working Group to continue to explore more advanced stock assessment models to make optimal use of the available data. The following sources of uncertainty should be addressed in subsequent assessments:
a ) The ICA calibration is based on a short survey series, and the calibration relationships are weak, with some year effects. The resulting estimates of terminal SSB are imprecise, and much of the uncertainty in terminal SSB is not included in estimates of precision. For example, a large portion of the spawning stock is composed of 1-ringers, for which proportion mature is poorly understood, and geometric mean abundance is assumed in the terminal year.
b) A stock assessment model that relaxes the assumption of constant selectivity in recent years should be explored. The separability assumption (and the assumed selectivity at the oldest ages for the entire time series) may lead to misinterpretation of the apparent shifts in age selectivity by the fishery. According to the 2009 HAWG report (page 298), the 2007 benchmark concluded that changes in
fishing pattern (and conflicting signals) prevented a final assessment from being conducted
c ) The magnitude of discarded catch should be estimated and included in the stock assessment.
d) A stock assessment approach that accounts for the mixed-stock resource and connectivity with adjacent management units should be developed.
e) Fishing mortality reference points (Flim and Fpa) should be proposed.
f) Consumption of Celtic Sea herring should be estimated and considered for stock assessment and fishery management.
2 ) The settings of the short-term forecast appear to be chosen correctly, and the $B_{l i m}$ and $B_{p a}$ reference points (as revised for retrospective change) are still appropriate. However, we suggest that the same forecast approach be extended for long-term, stochastic projection to determine the fishing mortality rate associated with $B_{\lim }$ (as a candidate for $\mathrm{F}_{\mathrm{lim}}$ ) and its uncertainty (to derive $\mathrm{F}_{\mathrm{pa}}$ and potentially a revised $\mathrm{B}_{\mathrm{pa}}$ ) as well as MSY reference points.
3 ) The methodology is generally well-described in the stock annex and allows repeatability. The various tables and information in the body of the report give the strong impression that all inputs are sufficiently documented to allow an outsider to do an assessment. However, some details of the most recent application of the stock assessment model (e.g., input and output tables, model diagnostics) should be provided to justify the modeling decisions.

## Detailed Comments (organized by Annex section)

## A. General

A.1. Stock definition - Several aspects of stock definition are described to justify the appropriateness of the management unit and identify aspects of population structure that may influence stock assessment and fishery management. Atlantic herring are 'population-rich' throughout their range, with complicated patterns of ontogenetic movement and mixing of spawning groups. These complex patterns present challenges to conventional stock assessment and fishery management.

1 ) Combined assessment of autumn and winter spawning groups appears to be the most appropriate use of available data, because of extensive mixing of spawning groups resulting in mixed-group fisheries and surveys. However, continued advancements in discrimination of seasonal spawning groups should be explored with the ultimate goal of stock composition analysis and consideration of spawning groups in assessment and management.
2 ) The inclusion of area VIIj in the Celtic Sea management unit appears to be appropriate because of similar demographic patterns in VIIj and g, larval mixing between the two areas, and a common nursery area in VIIj shared by herring spawned in VIIg. However, Figure 1 suggests that spawning in VIIg is primarily in winter, while spawning in VIIj occurs in both autumn and winter. Similar to the comment above, the seasonal spawning pattern
suggests that the development of stock composition analysis would facilitate the consideration of spawning groups in assessment and management.
3 ) The boundary between the Celtic Sea and Irish Sea herring management units is supported by the results of an extensive multidisciplinary program (WESTHER). Although the $52^{\circ} 30^{\prime} \mathrm{N}$ boundary is well-justified, advection of larvae from the Celtic Sea to the Irish Sea and subsequent return to spawn in the Celtic Sea has consequences to assessment and management of both resources. Loss of larvae from the Celtic Sea will add noise to the stock-recruit relationship. Depending on which age fish return to the Celtic Sea, the immigration may confound inferences of mortality from the catch-at-age analysis which assumes a closed population. Return migration at 1-ringers will have less influence on the perceived population dynamics than on immigration of older ages. The relationship between Celtic Sea and Irish Sea herring should be further investigated to better understand the sensitivity of the closed-population assumption in the assessment.

4 ) Similar to the comment above, the relationship between Celtic Sea herring and those in in VIIe-f and VIIIa should be investigated to understand the sensitivity of the closed-population assumption in the assessment.
5 ) Figures 1 and 2 are switched.
6 ) A Figure of the region should be provided that includes all of the areas described in the Annex (VIIe, VIIh, VIIk, VIIIa).
7 ) The species name Clupea harengus should be included in the Annex.

## A.2. Fishery

1. The fishery description is informative and well-written.
2. The increased landings after World War II support the premise that fishing influences stock size - a principle that should not be taken for granted for small pelagic species.
3. Any information on historical landings (prior to 1958) would be informative.
4. The statement that "Further fluctuations in the landings were evident during this time with high quantities of herring landed from 1958-1960 and from 19661971 (Molloy, 1972)" is somewhat inconsistent with the data plotted in Figure 6, in which annual landings from 1959 to 1963 are similar (i.e., the 'high quantities from 1958-1960 persisted to 1963).
5. The 'polyvalent' category of vessels should be described as in the 2009 HAWG report ("The term 'Polyvalent' refers to a segment of the Irish fleet, entitled to fish for any species to catch a variety of species, under Irish law" page 296).
6. The catch of large, old fish appears to depend on what areas are seasonally open to the fleet (e.g., Labadie Bank being open in July 2003 led to older fish being caught as compared to openings later in the year). Therefore, two aspects of the assessment model (separability in the recent period and full selectivity of the oldest age in all years) may be inappropriate.
7. The last paragraph "the Irish Quota" is redundant almost word for word with the second paragraph of the "Fishery in recent years"

## A.3. Ecosystem aspects

1. Given the important role of herring in the ecosystem, more information is needed on consumption of Celtic Sea herring and predation of other species by herring. Although estimates of herring consumption in the North Sea are used to derive the assumed natural mortality rates for Celtic Sea herring, consumption of herring in the North Sea should be estimated, and incorporation of consumption in stock assessment and management should be considered.
2. References should be provided for the statement that "studies have been unable to demonstrate that changes in the environmental regime in the Celtic Sea have had any effect on productivity of this stock."
3. The recent reduction in size-at-age should be reported in this section as a possible response to ecosystem factors. The trend is critical for estimation of spawning biomass, and the cause of the trend is important for assessment decisions and modeling future expectations (e.g., forecasts and reference points).
4. Similarly, the increasing recent trend in total mortality estimates from catch curves or $\log$ catch ratios and the decreasing recent trend in fishing mortality from the stock assessment model suggest an increase in natural mortality, which may reflect ecosystem change.
5. The ecosystem description and the summary of spawning dynamics suggest that there is adequate information to develop a bio-physical model of larval transport that would provide a complementary perspective on connectivity among spawning groups.
6. Differences in survival between the Irish and Celtic Seas could have important consequences to population dynamics. The relative contribution of each habitat to the adult population in the Celtic Sea should be determined.
7. Given that the Celtic Sea is near the southern extent of the range of herring, and the increasing trend in temperature, the potential for a northward shift in distribution should be monitored.

## Discards

1 ) The "discards" section of this part should probably be in the data section.
2 ) Although the discard rates are considered to be low and discards are not included in the assessment, discards occur. The report indicates that discarding is influenced by market situations, which suggests it may rise. While, it does not seem to have at the moment a potential impact on the assessment, it could be useful to evaluate the consequence of the inclusion of discards on an exploratory basis. The underestimate of total catch produces biased estimates of stock size and mortality. The discard rate estimated by Berrow, et al. (1998), $4.7 \%$, should be used to derive an approximate magnitude of discards that would be more accurate than the implicit assumption of no discards.
3 ) An at-sea monitoring program should be developed to estimate discard rates (including slippage) and to sample size and age structure of discards.
4 ) The statement that Berrow, et al. (1998) "indicated that the Celtic Sea herring fishery is very selective and that discard rates are well within the figures estimated for fishery models" is not clear. Does the statement suggest that a $4.7 \%$ discard rate is similar to that estimated for other fisheries, or similar to the rate assumed in fishery models? Celtic Sea herring stock as-
sessment assumes no discards and is not consistent with the estimate of a $4.7 \%$ discard rate.

## B. Data

## B.1. Commercial catch

1 ) Sampling intensity of the series of catch-at-age should be provided to evaluate the reliability of catch-at-age estimates. For example, Table 4.2.2.1 in the 2009 HAWG report indicates that 45 samples were collected from the 2008 fishery, and all major area-quarter components of the catch were sampled. Is this typical of the sampling intensity since 1958 or are there any systematic gaps in historical sampling that should be considered in the interpretation of catch-at-age? Given the complex pattern of time-area closures and fishing patterns, 45 samples per year may not adequately characterize some of the fine-scale differences in catch-at-age.
2 ) The report mentions the landings statistics for this stock need correction for misreporting. Landings apparently include substantial amounts ( $>10 \%$ ) of fish from other areas but while the correction is made (i.e. unallocated landings), no information is given on how this correction is done. This information is important to explain how to prepare the data from the raw landings statistics.
3 ) A requirement of logbook data for all vessels in the sentinel fishery could improve estimates of small boat landings.

## B.2. Biological

1 ) The various biological parameters are well described and their quality appears to be reliable for the assessment.
2 ) Have the age determination methods been validated?
3 ) Including some typical age-length keys would be helpful to evaluate how well catch-at-age is being estimated, particularly at older ages.
4 ) Including the quality-control results for precision estimates would be informative, and potentially useful for the development of advanced statistical catch-at-age models that use the pattern of disagreements to model errors in the catch-at-age.
5 ) The use of age-specific natural mortality rates from multispecies VPA is appropriate for a small, pelagic forage species, but a development of a MSVPA for the Celtic Sea would be more appropriate than using the results from the North Sea. A reference should be provided for the MSVPA so that its details do not need to be included in the Annex.
6 ) The cause of the reduction in weight-at-age should be explored further. More specifically, determining if it results from ecosystem factors or fishing patterns is essential for making the correct selectivity assumptions in the stock assessment. It would be valuable to inspect weight-at-age data from surveys to see if fishery-independent data reflect the same recent reduction. The beginning of the decline is consistent with the development of the roe fishery.
7 ) The choice of the maturity ogive suggests that various sources of information provide similar results. However, the amount of available information
(number of individuals sampled) is not included. The rationale for assuming $50 \%$ maturity of 1-ringers, rather than the estimated $58 \%$, is that the fishery probably samples precocious fish. However, the estimate of $58 \%$ is from a survey. The text does not report if the estimate of $>50 \%$ by Lynch (in prep.) is from the fishery or a survey. Given the substantial contribution of 1-ringers to the spawning stock, a more precise estimate of proportion mature should be applied.
8 ) Recruitment from the Irish sea may affect maturity of the population (and consequently maturity ogives). The possible influence of individuals from the Irish Sea raises the question of the proportion of individuals from that area and the effects of possible changes in maturity from one area to another. For some other stocks, like the Celtic Sea cod, a similar situation is observed and the lack of samples makes any maturity ogive rather uncertain. This was one of the criticisms in the benchmark review of the Celtic Sea cod.

## B.3. Surveys

1 ) The timing of the survey appears to be related to 'year effects' in calibration diagnostics (HAWG 2009 Figure 4.6.1.1), suggesting that the portion of the resource in the survey area is sensitive to the time of the survey. The 2002/2003 survey (conducted in September and October) has all positive residuals (i.e., more fish in the survey than predicted by the model), and the 2003/2004 survey (conducted in October and November) has all negative residuals (i.e., fewer fish in the survey than predicted by the model), suggesting that fewer fish are available to the survey later in the year. Is it possible that spawners are in the process of leaving the survey area during the survey?

2 ) Imprecision of survey estimates is illustrated by the large difference between estimates of 2008 SSB from the survey (90kt) and the assessment model (56kt).
3 ) For the acoustic survey, the estimates of CV appear to be based on a simple function of the positive number of samples. Some explanations about that relationship would have been welcome. An apparent contradiction is in the text: "CV was obtained based on transect mean densities," but mean density is not included in the equation. There appears to be something missing from the equation.
4 ) The decision to use a shorter, standardized series for a tuning index is valid.

5 ) More information is needed to describe how indices of abundance at age are derived from the acoustic survey.
6 ) Tuning is based only on the acoustic survey which apparently provides the best indices. Data from other surveys are not used. Some other stocks, like the Celtic Sea cod, have the same issues of having surveys that are not specifically targeting those species sampling few and variable numbers of fish. Some work has been carried out during the WKROUND benchmark to combine survey indices and some others stocks (e.g. Sole in VIId) use combined survey indices. The report and stock annex do not mention any attempt to use or combine the available information. Some exploratory work on using those datasets would also be welcome. This could involve evalu-
ating how the indices behave against each other, against fishing vessels and how they could affect the assessment. Some analysis of the trends of all survey data would be helpful to support the choice of only using the acoustic survey for the assessment.
7 ) The analysis of productivity over time is sufficiently commented to naturally end with the conclusion that recent F has been detrimental to the stock productivity.
8 ) If the Irish Groundfish Survey is expected to provide qualitative information for the assessment, results should be included in the Annex.
9 ) Similarly, if the Northern Ireland GFS survey offers a potential recruitment index, more details are needed in the Annex so that it can be considered as more information becomes available on natal origin.
10 ) Similarly, data from larval surveys should be provided as a comparison to stock assessment results.

## B.4. Commercial CPUE

The decision to exclude fishery CPUE as a tuning index in the stock assessment is valid, because of the nature of herring behavior, fishing patterns and management changes. However, it would be informative to compare the acoustic survey index to CPUE information from the fishing vessels. Some stocks (e.g. whiting in the North Sea) have conflicting patterns between surveys and fishing vessels. One reason could be some slight changes in the survey interfering with the results. Therefore, this type of comparison can be helpful to evaluate the consistency of the observation from the surveys in addition to the quality (i.e. level of noise) of the data which is another aspect to consider.

## C. Historical Stock Development

1) This section shows some issues with noisy data but does not seem to explain the "conflicting signals in input data and changes in the fishing pattern" referenced in the HAWG report from the 2007 benchmark.
2 ) The description of 'time periods in the fishery' is informative and suggests that some of the selectivity assumptions in the stock assessment should be reconsidered. The roe fishery targeted older, mature fish, which would lead to greater selectivity of the oldest age during that period.
3 ) Tables 3 and 4 referenced in 'Time Periods in the Fishery' should be Tables 4 and 5.
4 ) Estimates of total mortality from log catch ratios and catch surveys are informative, but the age ranges selected for catch-curve analysis are inconsistent with results from the stock assessment model, because of the assumed pattern of natural mortality at age and estimated selectivity at age. According to the assessment model, herring are not fully selected until age-3, and natural mortality of ages 2 and 3 is greater than than for ages $4+$, so catch curves should be revised from ages 2-7 to ages 4-7.
5 ) The increasing recent trend in total mortality estimates from catch curves or log catch ratios appear to contradict the decreasing recent trend in fishing mortality from the stock assessment model. The cause of the discrepancy (e.g., increasing natural mortality) should be explored.

6 ) In the assessment section, the HAWG report (4.6.1) mentions "conflicting signals in input data and changes in the fishing pattern" but no information is given on the "historic" choices made for exploratory assessments. The decisions about model configuration are well explained and the assessment well commented (changes in plus groups, shortening time series, terminal selection and reducing the separable period) providing the rationale for the parameters used in the final assessment.
7 ) The concern raised by the 2007 benchmark assessment about violating the assumed constant separable pattern was not addressed. The assessment model still assumes separability in recent years. We reiterate the concern about assuming constant selectivity and repeat the recommendation to consider alternative modeling approaches that relax this assumption.
8 ) The revision of the catch-at-age used for the stock assessment model (truncation to ages $1-6+$ ) produces a more realistic selectivity pattern than the previous analysis of ages 1-7+.
9 ) Diagnostic features of the stock assessment model are needed in the Annex to evaluate model performance. Standard diagnostics from the 2009 assessment should be included (e.g., HAWG 2009 Figures 4.6.1.2-3 model residuals, 4.6.1.4-5 confidence intervals, 4.6.1.6. historical comparisons, 4.6.2.1 calibration plots, 4.6.2.5-6 retrospective analysis). Inspection of diagnostics suggests that the ICA model is relatively consistent and has no strong patterns in catch residuals, but there are strong 'year effects' in the survey residuals (i.e., same direction of deviation at all ages) for the first three years of the six surveys used (Figure 4.6.1.1), and calibration relationships are relatively weak (Figure 4.6.2.1). A 'year effect' in the terminal year will present problems for estimating terminal abundance and determining stock status.
10 ) The advantage of ICA over other models is well explained as well as the reasons for adopting new parameters. Considering the few changes in the list of parameters, it can be confusing to have the parameters listed for the former and new assessment methods separately. Maybe combining both sets of parameters into a single table would be more useful (considering only 2 parameters of 8 changed) so no "quick reader" may switch to the wrong set of parameters.
11 ) The Annex reports that ICA was chosen because of its emphasis on young ages and greater consistency, but there is no information in the Annex or the 2009 HAWG report on the performance of alternative models. It is difficult to judge the validity of that conclusion without example results from all viable models. For example, did XSA also have year effects in survey residuals?
12 ) The ICA model appears to perform well for this application, but the method is somewhat dated (it is a re-codification of the CAGEAN model developed by Deriso et al. 1985). Catch-at-age models have evolved since the 1980s, and more advanced methods (e.g., statistical catch-at-age, SCAA) may be more appropriate for assessing the data available for Celtic Sea herring. SCAA would also be able to use all recent and historical information available (e.g., selectivity for each fishery and each period, calibration of historical abundance with discontinued surveys, discard rate estimates)

13 ) References for ICA (Patterson 1998) and FLICA (flr-project.org) are needed.

14 ) The analysis of productivity over time supports the conclusion that recent F has been detrimental to the stock productivity. However, the calculation of surplus production is either poorly described or inaccurate. The equation $\mathrm{Ps}=\mathrm{Br}+\mathrm{Bg}-\mathrm{M}$ does not account for the different units of biomass ( t ) and natural mortality rate $\left(\mathrm{y}^{-1}\right)$. The inappropriate mix of instantaneous rates and biomass is continued in the subsequent statement that net production is calculated as Ps - F. Surplus production should be calculated as $\mathrm{Ps}=\mathrm{Br}+\mathrm{Bg}-\mathrm{Bm}$ where Bm is biomass of fish that die of natural causes. Total Production should be calculated as $\mathrm{P}=\mathrm{Br}+\mathrm{Bg}-\mathrm{Bm}+\mathrm{Y}$ where Y is catch biomass.

## D. Short-Term Projection

1 ) This section does not explain why the MFDP projection was not carried out from 2005 to 2008 and why this analysis is back in the assessment. It seems that the information available is of sufficient quality to allow this type of projection.
2 ) The projection methodology is appropriate, but stochastic projection, incorporating uncertainty in abundance at age estimates and recruitment estimates would help to evaluate alternative management actions by providing probability of achieving management objectives or risk of exceeding limits.

## E. Medium-Term Projections

The text states that Fmsy is provided in Table 7, but it is not (nor can it be from a simple yield-per-recruit analysis).

## F. Long-Term Projections and G. Biological Reference Points

1 ) The reference points have not been revised. However, considering the changes in the assessment methodology and some evidence Blim should be revised upwards. Some work to investigate a possible change of Blim should be encouraged (or an explanation is required to explain why these reference points should be kept as they are).
2 ) Long-term, stochastic projection should be used to determine the fishing mortality rate associated with $\mathrm{B}_{\lim }$ (as a candidate for $\mathrm{Flim}_{\text {) }}$ ) and its uncertainty (to derive $\mathrm{F}_{\mathrm{pa}}$ and potentially a revised $\mathrm{B}_{\mathrm{pa}}$ ) as well as MSY reference points.

## H. Management and ICES Advice

1) The ICA model suggests a recent increase in spawning biomass (2008 SSB=55 800t) to greater than Bpa (44000t), such that a rebuilding program is no longer necessary. However, important caveats should be communicated in the management advice. The ICA calibration is based on a short survey series, and the calibration relationships are weak, with some year effects. The resulting estimates of terminal SSB are imprecise, and much of the uncertainty in terminal SSB is not included in estimates of precision. For example, a large portion of the spawning stock is composed of 1-
ringers, for which proportion mature is poorly understood, and geometric mean abundance is assumed in the terminal year.
2 ) Comparison of fishery yields and TACs indicates that the management system can effectively control the fishery (e.g., TAC was slightly exceeded in two years in the last 20 years).
3 ) Previous ICES advice that 'catches of around 5000 t would be associated with stock recovery' appears to be unsubstantiated, because catches have not been that low in the observed catch series.

## Annex 14 Technical Minutes of the North Sea ecosystem Review Group

Review of ICES HAWG Report 2009

| Reviewers: | Gary Melvin (Canada, chair) |
| :--- | :--- |
|  | Outi Heikinheimo (Finland) |
| Chair WG: | Norman Graham (Ireland) |
| Secretariat: | Tomas Gröhsler |
|  | Barbara Schoute |

## General

This HAWG was one of 3 working groups reviewed by the North Sea Technical group. The RG acknowledges the intense effort expended by the working group to produce the report and the work required to complete their documentation in a timely manner.

The Review Group considered the following stocks:

| her-3a22 | Herring in Division IIIa and Subdivisions 22-24 (Western Baltic spring spawners) |
| :--- | :--- |
| her-47d3 | Herring in Subarea IV and Divisions IIIa and VIId (North Sea autumn spawners) |
| spr-nsea | Sprat in Subarea IV (North Sea) |

Stocks which may need a benchmark in future are:

The North Sea Sprat in Subarea IV will be the subject of a benchmark assessment in September 2009

## Herring in Division IIIa and Subdivisions 22 - 24 (Western Baltic spring spawners) her-3a22

1 ) Assessment type: update
2 ) Assessment: analytical
3 ) Forecast: Short term forecast presented
4 ) Assessment model: FLICA
5 ) Consistency: Retrospective analysis: Bias (20\%) in SSB and F compared to previous assessment.
6 ) Stock status: SSB and F stable, decreasing trend in recruitment since 2003. Recruitment the lowest observed. F is larger than any proxy of $\mathrm{F}_{\text {msy. }}$ SSB $2008=159,406,2009=141,824, \quad F 3-6=0.37 .2008$ recruitment estimated to be lowest in last 18 years
7 ) Man. Plan.: At an early stage. No defined reference points. However, exploratory management plan of $\mathrm{Fmsy}=0.25$, Blim=110,00, TAC variation $15 \%$, and Target $\mathrm{F}=0.25$

## General comments

The spring spawning stock is composed of several rather distinct spring spawning populations and probably different sub-stocks, and partly mix with the North Sea autumn spawners in IIIa and IVa(east). In addition, the stock is exploited by fleets from several countries. This makes the assessment and management planning extremely complicated. A substantial part of the catch reported as taken in Division IIIa by fleet $C$ was actually has been taken in Subarea IV

Fleet C - Directed fishery by trawlers and purse seiners, Fleet D - all trawlers and small purse seiners (Danish and Swedish) fishing for sprat, Norway pout and blue whiting. And Fleet-F (SD 22-24) in Western Baltic

The 2003 year-class has been largest component of the SSB for last 3 years.
According to the WG, the overall sampling in 2008 more than meets the recommended level and the coverage of areas, times of the year and gear (mesh size) was acceptable. Only in Subdivision 23 is the temporal coverage not acceptable. Discards considered insignificant.

The maturity ogive of WBSS applied in HAWG has been assumed constant between years although large year-to-year variations in the percentage mature have been observed. A Workshop on Sexual Maturity Staging of Herring and Sprat is taking place during 2009 in order to establish correspondence between old and new scales to convert time series and propose optimal sampling strategy to estimate accurate maturity ogives.

## Technical comments

3.6.1.2.: The estimates of natural mortality were derived as a mean for the years 1977-1995 from the Baltic MSVPA (ICES 1997/J:2). Would more recent estimates be available, and would there be any difference?
3.11: Recent recruitment has dropped appreciably and consistently, while stock size has remained constant. This indicates an environmental effect, which is not discussed in "Ecosystem considerations". (This issue is discussed in the report for NSAS). Are
there attempts to examine this phenomenon and take it into account in the management planning? The target $\mathrm{F}=0.25$ may be too high if the reproductive success is low and leads to a decrease of the spawning stock in coming years.

## Conclusions

The assessment is correct but more profound ecosystem considerations would be necessary.

## Herring in Subarea IV and Divisions IIIa and VIId (North Sea autumn spawners) her-47d3

1 ) Assessment type: update
2 ) Assessment: analytical
3 ) Forecast: Short-term forecast presented.
4 ) Assessment model: FLICA. 4 indices; Acoustic, Larval, IBTS, and MIL(0ringer)
5 ) Consistency: Retrospective analysis: consistency good, exception recruitment, age 0 in 2008.
6 ) Stock status: SSB above Blim but increased risk (below Bpa), recruitment better than in recent years but still low, F appropriate but above target.
7) $2008 \mathrm{SSB}=1.0$ million t , Bpa ( 1.3 million t ), Blim ( 800,000 ) with trigger at $\mathrm{B}=1.5$ million. $\mathrm{F}_{2-6}$ in $2008=0.24$, target $\mathrm{F}_{2-6}=0.14$. The 2008 year-class is higher than the previous 6 years but similar to the 2001 year class
8 ) Man. Plan.: Yes, adjusted in November 2008. Currently reduction of F advised. $\mathrm{F}_{2-6}(0.24)$, in 2008 above the target $\mathrm{F}_{2-6}$ of 0.14 . SSB is not expected to reach Bpa in 2010 even without fishery, but it may reach Bpa in 2011 with a substantial reduction (well over $30 \%$ ) in catches.

## General comments

The sampling coverage has decreased from 2007 but the spread of the effort over the different métiers was a little better than 2007. Information on discards has improved but is still on a low level.

There was an increase in recruitment from last year's estimate, which was outstandingly low, but was one in a series of poor recruitments starting from the 2002 year class.

North Sea herring is nominally being managed by a precautionary management plan, although the SSB is now below the precautionary biomass reference point. HAWG considers that the parameters of the management plan should take primacy over the management against precautionary reference points Fpa or Bpa. Not following the management plan has resulted in the SSB being at greater risk of being below $\mathrm{B}_{\mathrm{lim}}$ and lower catches

The low reproductive success is discussed in Environmental considerations (2.13). An ICES study group has reviewed the hypotheses for the serial poor recruitment in North Sea herring. Further investigation of the causes of the poor recruitment will require targeted research projects.

## Technical comments

The report might be easier to read when the text would be in the beginning and all large tables and figures after the text (as in the WBSS assessment report).

All predictions are for North Sea autumn spawning herring only.

## Conclusions

The SSB is expected to increase slightly both in 2010 and further in 2011.
The assessment has been performed correctly and the RG agrees with the conclusions

## Sprat in Subarea IV (North Sea) spr-nsea

1) Assessment type: update
2) Assessment: not presented
3) Forecast: not presented
4) Assessment model: no assessment was performed this year, see below
5) Consistency:
6) Stock status: Abundance, recruitment and catches at a low level
7) Man. Plan.: No. Management by TAC, ITQ

## General comments

Previous exploratory assessments of this stock have been performed using the CSA method, and catch prediction for the assessment year was provided on the basis of a linear regression of catch (as estimated by landings) versus the IBTS sprat index.

Boot strapping indicated the upper confidence limit ranges from $30 \%$ to $4600 \%$ greater than the index estimated by ICES, with a median value of $250 \%$. The lower confidence limit ranges from $20 \%$ to $90 \%$ less than the value index, with a median value of $40 \%$ However, these methods were found clearly inappropriate for the task at hand, and the results meaningless in an advice context. The decision was therefore made by HAWG not to perform or report any such runs this year.

The 2005 index (2004 year class) was the highest for the time series prior to 2009. The incoming 1-group (2008 year class) is estimated to be the highest for the whole time series, both in absolute and relative terms but this estimate should be considered as preliminary
This stock will be the subject of a benchmark assessment in September 2009, as part of the WKSHORT workshop.

## Technical comments

The IBTS results are presented in Table 8.3.1. and Figures 8.3.1. a-c. How reliable are these results? Uncertainty about these estimated used as reasoning for not presenting an assessment or a forecast. There could be a footnote to remind the reader that the indices for some years are highly uncertain.

Fig. 8.3.5. cannot be interpreted because of many overlapping curves. Some examples could be presented only.

## Conclusions

The decision not to present any assessment or forecast because of large uncertainties was justified.


[^0]:    * majority of catches landed to ljmuiden, the Netherlands

[^1]:    *incl. mean for Sub-division 23, which was not covered by RV SOLEA
    **incl. mean for Sub-division 21, which was not covered by RV SOLEA

[^2]:    *Geometric mean 1970-2007.

[^3]:    *Geometric mean 1970-2007.

[^4]:    * no information, but catch is likely to be negligible

[^5]:    ${ }^{1}$ sprat only; ${ }^{2}$ Data can be made available for the IoM waters only

[^6]:    *Average for the preceding nine years.

[^7]:    * Preliminary

[^8]:    ${ }^{1}$ http://www.gma.org/herring/biology/life_cycle/default.asp

[^9]:    * Average for the preceding five years

[^10]:    a. Unusually large catch removed, b. unusually large catch retained.

