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## i Executive summary

The Working Group for the Celtic Seas Ecoregion (WGCSE) performs stock assessments on demersal stocks in Rockall, West of Scotland, Irish Sea, West of Ireland, Western English Channel, Bristol Channel, Celtic Sea and Southwest of Ireland. The working group provides updated fisheries data and reviews for ten Nephrops stocks, four sole and plaice stocks, three cod and whiting stocks, three haddock stocks, two megrim, one seabass, one anglerfish and one pollack stock. For most of the stocks, advice is drafted in May for June release. Advice for Nephrops, anglerfish and Rockall megrim are not issued until autumn to take account of the 2021 survey information.
For a number of stocks (bss.27.6a7bj, cod.27.6b, ple.27.7bc, sol.27.7bc, whg.27.7a, nop.27.6a, whg.27.6b, nep.27.6aoutFU, nep.27.7outFU) no new advice was provided this year.
Two stocks have gone through a benchmark procedure in the past year; cod.27.7a, ple.27.7f-g. the results of which were presented to the group. Analytical assessments using age-structured models were conducted for 12 of the fish stocks. Surplus-production models, without age or length structure, were used to assess lez.27.4a6a and lez.27.6b, and a Depletion-Corrected Average Catch model to assess pol.27.67.
In 2022 the state of the five fish stocks for which no analytical assessment could be performed were inferred from application of Data-Limited Methods, using survey or biomass indices as indicators of stock development along with indications of stock status inferred from length indicators.
UWTV survey-based assessments were conducted for ten Nephrops stocks. Overall the stock status across the ecoregion show a decline in abundance to last year, with a reduction of circa 8800 tonnes in finfish advice (mostly owing to reduction in Celtic sea haddock [had.27.7.bk]; down 4,045 tonnes, and Celtic sea whiting [whg.27.7.b,c,e-k]; down 2737 tonnes), in the order of a $14 \%$ reduction on 2021 advice. Of the 22 assessed fish stocks, five stocks were fished above FmSY, ten below and seven stocks had unknown status relative to Fms;; 8 were above MSY Blim, and seven below $B_{l i m}$, with the status of seven unknown.
Of the eleven Nephrops stocks, catch advice saw an increase of circa 2800 tonnes on 2021 advice (approximately 8\%), with a range of Functional Unit increases and decreases.

## ii Expert group information

| Expert group name | Working Group for the Celtic Seas Ecoregion (WGCSE) |
| :--- | :--- |
| Expert group cycle | Annual |
| Year cycle started | 2022 |
| Reporting year in cycle | $1 / 1$ |
| Chairs | Jonathan White, Ireland |
| Meeting venues and dates | 4-13 May 2022, Online meeting, 29 participants <br> 13-15 September 2022, Online meeting, 15 participants |

## 1 Introduction

The Introductory section will be completed in autumn when the WGCSE 2021 report is finalised.

### 1.1 Terms of reference

### 1.1.1 Generic ToRs for Regional and Species Working Groups

2021/2/FRSG01 The following ToRs apply to: AFWG, HAWG, NWWG, NIPAG, WGWIDE, WGBAST, WGBFAS, WGNSSK, WGCSE, WGDEEP, WGBIE, WGEEL, WGEF, WGHANSA and WGNAS.

The working group should focus on:
a) Consider and comment on Ecosystem and Fisheries overviews where available;
b) For the aim of providing input for the Fisheries Overviews, consider and comment on the following for the fisheries relevant to the working group:
i) descriptions of ecosystem impacts on fisheries;
ii) descriptions of developments and recent changes to the fisheries;
iii) mixed fisheries considerations; and
iv) emerging issues of relevance for management of the fisheries;
c) Conduct an assessment on the stock(s) to be addressed in 2022 using the method (assessment, forecast or trends indicators) as described in the stock annex; - complete and document an audit of the calculations and results; and produce a brief report of the work carried out regarding the stock, providing summaries of the following where relevant:
i) Input data and examination of data quality; in the event of missing or inconsistent survey or catch information refer to the ACOM document for dealing with COVID19 pandemic disruption and the linked template that formulates how deviations from the stock annex are to be reported.
ii) Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information.
iii) For relevant stocks (i.e. all stocks with catches in the NEAFC Regulatory Area), estimate the percentage of the total catch that has been taken in the NEAFC Regulatory Area in 2021.
iv) For category 3 and 4 stocks requiring new advice in 2022, implement the methods recommended by WKLIFE X (e.g. SPiCT, rfb, chr, rb rules) to replace the former 2 over 3 advice rule ( 2 over 5 for elasmobranchs). MSY reference points or proxies for the category 3 and 4 stocks.
v) Evaluate spawning-stock biomass, total stock biomass, fishing mortality, catches (projected landings and discards) using the method described in the stock annex;

1. for category 1 and 2 stocks, in addition to the other relevant model diagnostics, the recommendations and decision tree formulated by WKFORBIAS
(see Annex 2 of https://www.ices.dk/sites/pub/Publication\ Reports/Ex-pert\ Group\ Report/Fisheries\ Resources\ Steer-
ing\%20Group/2020/WKFORBIAS 2019.pdf) should be considered as guidance to determine whether an assessment remains sufficiently robust for providing advice.
2. If the assessment is deemed no longer suitable as basis for advice, consider whether it is possible and feasible to resolve the issue through an interbenchmark. If this is not possible, consider providing advice using an appropriate Category 2 to 5 approach;
vi) The state of the stocks against relevant reference points;

Consistent with ACOMs 2020 decision, the basis for $\mathrm{F}_{\mathrm{pa}}$ should be $\mathrm{F}_{\mathrm{p} .05}$.

1. Where Fp. 05 for the current set of reference points is reported in the relevant benchmark report, replace the value and basis of Fpa with the information relevant for Fp. 05.
2. Where Fp. 05 for the current set of reference points is not reported in the relevant benchmark report, compute the Fp. 05 that is consistent with the current set of reference points and use as Fpa. A review/audit of the computations will be organized.
3. Where Fp. 05 for the current set of reference points is not reported and cannot be computed, retain the existing basis for Fpa.
vii) Catch scenarios for the year(s) beyond the terminal year of the data for the stocks for which ICES has been requested to provide advice on fishing opportunities;
viii)Historical and analytical performance of the assessment and catch options with a succinct description of associated quality issues. For the analytical performance of category 1 and 2 age-structured assessments, report the mean Mohn's rho (assessment retrospective bias analysis) values for time-series of recruitment, spawningstock biomass, and fishing mortality rate. The WG report should include a plot of this retrospective analysis. The values should be calculated in accordance with the "Guidance for completing ToR viii) of the Generic ToRs for Regional and Species Working Groups - Retrospective bias in assessment" and reported using the ICES application for this purpose.
d) Produce a first draft of the advice on the stocks under considerations according to ACOM guidelines.
i. In the section 'Basis for the assessment' under input data match the survey names with the relevant "SurveyCode" listed ICES survey naming convention (restricted access) and add the "SurveyCode" to the advice sheet.
e) Review progress on benchmark issues and processes of relevance to the Expert Group.
i. update the benchmark issues lists for the individual stocks in SID;
ii. review progress on benchmark issues and identify potential benchmarks to be initiated in 2023 for conclusion in 2024;
iii. determine the prioritization score for benchmarks proposed for 2023-2024;
iv. as necessary, document generic issues to be addressed by the Benchmark Oversight Group (BOG).
f) Prepare the data calls for the next year's update assessment and for planned data evaluation workshops;
g) Identify research needs of relevance to the work of the Expert Group.
h) Review and update information regarding operational issues and research priorities on the Fisheries Resources Steering Group SharePoint site.
i) If not completed in 2020, complete the audit spread sheet 'Monitor and alert for changes in ecosystem/fisheries productivity' for the new assessments and data used for the stocks. Also note in the benchmark report how productivity, species interactions, habitat and distributional changes, including those related to climate-change, could be considered in the advice.

Information of the stocks to be considered by each Expert Group is available here.

### 1.1.2 Specific ToRs

2021/2/FRSG10 The Working Group for the Celtic Seas Ecoregion (WGCSE), chaired by Mathieu Lundy, UK and Jonathan White*, Ireland will meet virtually 3-13 May 2022 and by correspondence September / October 2022 to:
a) Address generic ToRs for Regional and Species Working Groups;

The assessments will be carried out on the basis of the stock annex. The assessments must be available for audit on the first day of the meeting.

Material and data relevant for the meeting must be available to the group on the dates specified in the 2021 ICES data call.

WGCSE will report by 25 May 2022 for the attention of ACOM, and by 1 October 2022 for Nephrops stocks, anglerfish and megrim in Rockall.

Only experts appointed by national Delegates or appointed in consultation with the national Delegates of the expert's country can attend this Expert Group.

### 1.2 Participation

Due to the COVID-19 pandemic and uncertainties in Europe owing to the Russian invasion of Ukraine, the decision was taken in mid-March to hold the meeting remotely in order to give participants certainty in the meeting process. An adequate participation was achieved with representation of the different institutes generally involved.

### 1.3 Methods

The type of final assessments presented at the WG are summarised as follows:
Category 1 age-based assessments and forecasts were conducted for bss.27.4bc7ad-h, cod.27.7.ek, had.27.7.a, had.27.7.b-k, had.27.6a, ple.27.7.a, sol.27.7.a, sol.27.7.e, sol.27.7.fg, whg.27.6.a and whg.27.7.b-ce-k; cod 27.7.a went through benchmark in 2022 (WGNSCS, ICES 2022), and is now assessed as a Category 1 assessment.
Category 1 Bayesian surplus production model for lez.27.4.a6.a;
Category 1: UWTV survey based assessments and advice will be used for nep.fu.11, nep.fu.12, nep.fu.13, nep.fu.14, nep.fu.15, nep.fu.16, nep.fu.17, nep.fu.19, nep.fu. 2021 and nep.fu.22. Fisheries data were updated at the May meeting and survey data were updated in the autumn;
Category 2: Lez.27.6b The stock has a SPICT assessment to determine stock status and a shortterm catch forecast;

Category 3: Several stocks are now assessed as Data-Limited following the guidelines of WKLIFEIX and X, (ICES, 2019; 2020) following the "rfb" approach, implementing trends from combined biomass index and length-based indicators as the basis for advice. These include: ple.27.7.e; ple.27.7fg; ple.27.7h-k with trends from combined biomass index. Further, assessed in the autumn were anf.27.3a46 and had.27.6.b,

Ple.27.7fg went through a benchmarked process in 2022 (WKNSCS ICES, 2022). The resulting assessment and basis for providing forecast catch advice was approved. During WGCSE however, a number of issues became apparent with the assessment from which it was decided to revert the advice basis to the data-limited, Category 3 " rfb " approach. These are detailed in the Report Section on ple. 27.7 fg , and may be summarised as:

- A lack of justification or objectivity in choice of natural mortality (m) levels.
- Smooth trend in Fishing pressure resulting from no correlation process implemented across $m$ age classes ( F -at-age lognormal random walks were not correlated across ages).
- When correlations were introduced they produced a less smooth F trend over time, and strong model retrospective patterns in $\mathrm{R}, \mathrm{SSB}$ and $\mathrm{F}(-0.19,0.53$ and -0.36 respectively) outside recommended limits.
- This also resulted in substantial changes in SSB and F in recent years.
- Recruitment in the assessment was modelled as a constant mean, removing this to enable recruitment to be calculated for each year from the data resulted in slight changes in SSB, F and R.

Category 4: Depletion corrected average catch was used for pol.27.67;
Category 5: sol.27.7h-k.
For the stocks for which a full analytical assessment was possible, the WG used either Extended Survivor's Analysis (XSA), Age-Structured Assessment Program (ASAP) or state-space assessment model (SAM). These approaches and procedures for using them are discussed in further detail in the relevant stock annexes.

### 1.4 Data issues

Data were generally submitted in a timely fashion through the InterCatch database for landings and discards data, and through the accessions database for other sources of data.

### 1.5 Transparent Assessment Framework (TAF)

TAF is a new framework, currently in development, to organize all ICES stock assessments. Using a standard sequence of $R$ scripts, it makes the data, analysis, and results available online, and documents how the data were pre-processed. Among the key benefits of this structured and open approach are improved quality assurance and peer review of ICES stock assessments. Furthermore, a fully scripted TAF assessment is easy to update and rerun later, with a new year of data. A number of assessments are being scripted in standard TAF scripts. See http://taf.ices.dk for more information and https://github.com/ices-taf/ for details.

### 1.6 Internal auditing and external reviews

As in previous years the WG carried out its own internal audit process using the standard ICES template. Given the workload of many of the scientists at WGCSE (sometimes with one scientist
responsible for two or more stocks), many of the reports were not finalized until after the WG meeting. Audits were therefore typically carried out by correspondence after the WG and not completed for some stocks. All stocks for which advice was provided in June and October 2022 were audited by the WG and audit reports were produced for most of these. Issues discovered during the audit process were corrected in the WG report.

### 1.7 Generic ToR e: WGCSE recommendations for stocks to be benchmarked

Stocks recommended for next round(s) of benchmarks:

| Listed for 2022-2023 Benchmark | Requested from 2023-2024 Benchmark |
| :--- | :--- |
| pol.27.67 | pok.27.7-10 |
| bss.27.4bc7ad-h | ple.27.7e |
| cod 27.6a | lez.27.4a-6a |
|  | ple.27.7fg |
| Requested for 2022-2023 Benchmark | sole 7a |
| had.27.6b | whg.27.7b-ce-k |
|  | lez.27.4a-6a |

pol.27.67 and pok.27.7-10 were recommended for benchmark in 2020.
Currently, pol.27.67 is categorized as category 4 data-limited and the DCAC method is applied to provide advice. As the DCAC method only uses long time-series of official landings, it may not reflect recent stock fluctuations or changes in the fisheries, smoothed by the length of the time-series. As a result, new computations of DCAC are always very close to the previous year's results even if recruitment or SSB highly fluctuate. Therefore, it is relevant to explore new assessment models. No assessment or advice has been provided for the pok.27.7-10 stock, as a benchmark should establish if the DCAC approach can be improved upon.
had.27.6b was recommended for benchmark in 2020.
At-sea observer sampling for discards remains sparse for had.27.6b, which leads to uncertainty in fishery selectivity patterns and catch estimates data used in the assessment. The assessment model used (FLXSA) assumes catch is measured with no uncertainty and so does not account for this sampling issue. The estimates of SSB are consistently being overestimated and F is consistently being underestimated, therefore it is recommended to address this in a benchmark.

## Atlantic seabass stocks were recommended for benchmark in 2020.

There was a joint recommendation with the Working Group for the Bay of Biscay and the Iberian Waters Ecoregion (WGBIE) to evaluate the stock identity of the Atlantic seabass stocks. The Working Groups recognized the complexity and considered that a stock identity workshop might be convened to allow relevant experts to consider relevant studies (data storage tags, conventional tags, genetics, otolith microchemistry and larval dispersion models) and advise whether the existing stock boundaries remain appropriate. This work should be proceeding towards a benchmark in the next stage. The aim should be to explore and peer-review all available information on recreational catches. There is also a need for information on recruitment trends in other areas, as it cannot be assumed that the Solent index will in the long-term represent overall recruitment patterns throughout areas 4 and 7 . Estimates of discards are available only from the early 2000s, but do not cover all fisheries, are imprecise, and are only included for some fleets in the assessment. Discard rates are expected to increase in the short term as fishers adjust to take account of the management measures, such as the increase in minimum conservation reference
size from 36 cm to 42 cm . The difference in perception between the modelled discards and the observed, should also be addressed. The benchmark should evaluate if sampling is currently sufficient to support continued application of Stock Synthesis fitting selection parameters to fishery composition data.

## In 2022 WGCSE recommend the following stocks to be benchmarked

ple.27.7e: A SAM assessment has been developed and is ready for critique and review.
whg.27.7b-ce-k: Following application in WGCSE the model has high retrospective patterns which are evidenced in Mohn's rho values from recruitment, SSB and fishing pressure. The assessment needs to be investigated and developed to improve its internal consistency and improve retrospective convergence.
lez.27.4a-6a: The present modelling approach is built on ageing IT architecture and more userfriendly applications are now available. A SPiCT model and SS3 should be explored as alternatives to the present labour-intensive approach.
ple.27.7fg: The SAM assessment developed in WKNSCS (ICES, 2022) should be reviewed, with qualification for natural mortality levels detailed and correlation linkages across age classes implemented to assess potential relative to retrospective patterns. A SPiCT model may also be appropriate.
had.27.6b: While catch and survey data are presently available their processing needs to be reviewed as the approached are currently not visible. For 2022 a survey based DLS approach was advised by WGCSE, while a winter-spring data review and model development benchmark would provide grounds to provide a Category 1 based assessment for 2023.

Every year a prioritization exercise for the stocks that need to be benchmarked is done. The sum of the weighting scores (1-5) for each of the five criteria will determine the urgency for a benchmark. Those criteria are related to the quality of the previous assessments, the opportunity to improve the assessment, the management importance, the perceived stock status and the time since the previous benchmark. To have an overview of this information, an issue list is requested for every stock.

# 2 Anglerfish (Lophius budegassa, Lophius piscatorius) in subareas 4 and 6 and in Division 3.a (North Sea, Rockall and West of Scotland, Skagerrak and Kattegat) 

## Assessment in 2022

The last benchmark for this stock was carried out in February 2018 (ICES, 2018) where it was agreed to provide advice on the basis of the procedure for category 3.2.0 of ICES RGLIFE datalimited stock (DLS) methods as set out in the stock annex. However, in 2022, based on the recommendations of the WKLIFE workshop, ICES decided that this method should no longer be used and future advice should be provided following the $r f b$ rule (ICES, 2021).

## ICES advice applicable to 2021 and 2022

## ICES advice for 2021

ICES advises that when the precautionary approach is applied, catches in 2021 should be no more than 17645 tonnes.

ICES advice for 2022
ICES advises that when the Precautionary approach is applied, catches in 2022 should be no more than 14116 tonnes.

### 2.1 General

## Stock description and management units

The anglerfish stock on the Northern Shelf is considered to occur in Division 3.a (Skagerrak and Kattegat), Subarea 4 (the North Sea) and Subarea 6 (West of Scotland plus Rockall). Anglerfish in the North Sea and Skagerrak/Kattegat were considered by this Working Group for the first time in 1999. In 2004 the WGNSDS considered the stock structure of anglerfish on a wider European scale, and found no conclusive evidence to indicate an extension of the stock area northwards to include Division 2.a. In 2013, Division 2.a was removed from WGCSE ToR.

## Management applicable to 2021 and 2022

Council Regulation (EU) 2021/92 of 28 January 2021 fixing for 2021 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters

| Species | Anglerfish <br> Lophiidae | Zone: | United Kingdom and Union waters of 4; United Kingdom waters of 2a (ANF/2AC4-C) |
| :---: | :---: | :---: | :---: |
| Belgium | $312(1)_{(1)}{ }^{2}$ |  |  |
| Denmark | $688\left({ }^{1}\right)\left({ }^{2}\right)$ |  |  |
| Germany | $336(1)_{(1)}{ }^{2}$ |  |  |
| France | $64\left({ }^{1}\right)\left({ }^{2}\right)$ |  |  |
| The Netherlands | $236(1)^{(2)}$ |  |  |
| Sweden | $8\left({ }^{1}\right)\left({ }^{2}\right)$ |  |  |
| United Kingdom | $10328\left({ }^{1}\left(^{2}\right)\right.$ |  |  |
| Union | $1644(1)^{(2)}$ |  |  |
| TAC | 11972 |  | Precautionary TAC |

(1) Special condition: of which up to $\mathbf{3 0 \%}$ may be fished in United Kingdom, Union and international waters of $\mathbf{6 a}$ north of $58^{\circ} 30^{\prime} \mathrm{N}$ (ANF/*6AN58).
(2) Special condition: of which up to $10 \%$ may be fished in United Kingdom waters of 6 a south of $58^{\circ} 30^{\prime} \mathrm{N}$;

United Kingdom and international waters of 5b; International waters of 12 and 14 (ANF/*56-14).

| Species | Anglerfish <br> Lophiidae | None: |
| :--- | :---: | :--- |
| (ANF/04-N.) |  |  |

(1) Special condition: of which up to $20 \%$ may be fished in: United Kingdom and Union waters of 2a and 4
(ANF/*2AC4C).

According to the 'agreed record', the following TACs have been agreed between the EU and UK for 2022. The allocations are as per the EU-UK Trade and Cooperation Agreement (TCA).

| ICES areas | TAC | UK allocation | EU allocation |
| :--- | :--- | :--- | :--- |
| UK \& EU waters of 4; 9014 $7849^{1,2}$ $1165^{1,2}$ <br> UK waters of 2a.    |  |  |  |
| 6; UK \& international <br> waters of 5b; interna- <br> tional waters of 12 \& 14 |  | $2060^{3}$ | $3042^{3}$ |

1) Special condition: of which up to $30 \%$ may be fished in United Kingdom, European Union and international waters of 6a north of $58^{\prime} 30^{\prime \prime}$ (ANF/*6AN58).
(2) Special condition: of which up to $10 \%$ may be fished in United Kingdom waters of 6a south of $58^{\prime} 30^{\prime \prime}$; European Union and international waters of 5b; International waters of 12 and 14 (ANF/*56-14)

3: Special condition: of which up to $20 \%$ may be fished in United Kingdom and European Union waters of 2a and 4 (ANF/*2AC4C).

| Species | Anglerfish <br> Lophiidae | Zone: | Norwegian waters of 4 <br> (ANF/04-N.) |
| :--- | :--- | :--- | :--- |
| Belgium | 37 | Precautionary TAC |  |

Management of Northern Shelf anglerfish is based on separate TACs for the North Sea Subarea 4 and West of Scotland Subarea 6. There is no TAC for Skagerrak and Kattegat Division 3.a. Table 4.1 summarises the ICES advice and actual management applicable for Northern Shelf anglerfish for 2003 onwards.

Although there is no minimum landing size for this species, there is an EU minimum weight of 500 g for marketing purposes (EC Regulation 2406/96).

## Fishery description

A more detailed description of the fisheries can be found in the Stock Annex. The official national landings as reported to ICES are given in Table 4.2 and the breakdown by country in Tables 4.34.5. Total officially reported landings of anglerfish from the Northern Shelf are shown in Figure 4.1.

Trends in nominal international fishing effort in the North Sea and Eastern Channel and the West of Scotland collated by STECF for the Evaluation of Fishing Effort Regimes in European Waters are shown in Figure 4.2. Since 2014, there have been slight increases in TR effort in both the North Sea and West of Scotland, with effort across all gears in the North Sea stable or reducing since 2012 and in the West of Scotland increasing from 2014 driven by marked increases in trawl fisheries. Data for 2017 have not yet been released by STECF although a significant change in the overall observed trend of anglerfish fleets is not anticipated with the introduction of 2017 data.

## The fishery in 2021

Official landings in 2021 for subareas 6 and 4 were 19154 t ( 5992 t and 13162 t respectively), giving a $17.5 \%$ undershoot of the combined TAC of 22056 t ( $81 \%$ and $83 \%$ TAC uptake respectively). In Subarea 6 Belgium (0\%), the Netherlands (0\%) and France (37\%) had noticeably low uptakes. Belgium was also observed to significantly undertake their quota in Subarea 4 ( $44 \%$ ). Denmark ( $58 \%$ ) and Germany ( $65 \%$ ) both decreased their Subarea 4 uptakes in comparison to 2019, while the United Kingdom increased its uptake (70\%). The UK exceeded its quota in Subarea 6 (by $52 \%$ ), a decrease of $24 \%$ compared to 2019 . Over quota landings by individual states are most likely due to countries obtaining additional quota from other EU member states, or carrying forward unutilised quota from 2019 and using a flexibility allowance whereby $10 \%$ of Subarea 4 TAC can be utilised to reattribute landings from Subarea 6.

Uptake of EC quota in 2021, based on the preliminary officially reported landings, was as follows:

|  | TAC 6 | Lan-dings 6 | Uptake (\%) | TAC 4 <br> (Norwegian) | $\begin{aligned} & \text { TAC 2.a \& } \\ & 4 \end{aligned}$ | TAC 2.a \& 4(total) | Landings <br> 4 | Uptake (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 202 | - | 0\% | 37 | 312 | 349 | 290 | 83\% |
| Denmark | - | - | - | 935 | 688 | 1623 | 1462 | 90\% |
| France | 2485 | 1326 | 53\% | - | 64 | 166 | 108 | 65\% |
| Germany | 230 | 157 | 68\% | 15 | 336 | 893 | 228 | 65\% |
| Ireland | 562 | 684 | 122\% | - | - | - | - | - |
| Netherlands | 194 | - | 0\% | 13 | 236 | 249 | 376 | 151\% |
| Norway | - | - | - | - | - | - | 982 | - |
| Russia | - | - | - | - | - | - | - | - |
| Spain | 216 | 282 | 131\% | - | - | - | - | - |
| Sweden | - | - | - | - | 8 | 8 | 113 | $1412 \%$ |
| UK (total) | 2488 | 3539 | 142\% | - | 10328 | 10328 | 9604 | 93\% |
| Total | 6377 | 6434 | 101\% | 1000 | 11972 | 13616 | 13162 | 97\% |

Based on data submitted to ICES, the fishery was principally prosecuted by vessels using demersal trawls (Table 4.6), targeting either white fish ( $83 \%$ of total landings by weight) or Nephrops $(5 \%)$. Alongside these fleets there was also a significant gillnet fishery ( $11 \%$ ), as well as an assortment of other gears in which small quantities of anglerfish are caught as bycatch. The latter have been grouped here as miscellaneous gears ( $1 \%$ ). Gillnets accounted for smaller proportion of total landings across gear types in 2021 in comparison to 2020.
UK (Scottish) vessels accounted for the majority of reported anglerfish landings from the combined Northern Shelf area, taking approximately $65 \%$ of the landings overall. Scottish, Danish and Norwegian vessels took $73 \%, 11 \%$ and $7 \%$, respectively, of the North Sea (Divisions 4.a-4.c) landings. Scottish, French and Irish vessels took $59 \%$, $22 \%$ and $11 \%$, respectively, of the West Coast (Subarea 6) landings.

Landings in Division 3.a are not regulated: Table 4.5 shows the official landings which fluctuated between 400-500 t from 2005-2015, but have more than doubled since then. Official landings in 2021 were 912 t , slightly higher than 2020.

### 2.2 Data

## Landings

National landings data as reported to ICES and Working Group estimates of total landings are given in Table 4.2. The working group procedures used to determine the total international landings numbers and weights-at-length are documented in the stock annex. It is acknowledged that throughout the landings time-series, there have consistently been differences between the total
official landings and the landings as estimated by the WG. This is likely due to differences in the data provided to the WG by national scientists and administrators.

Due to restrictive TACs, the likelihood of misreporting and underreporting of anglerfish landings in the past is considered to have been high, particularly during the period 2003-2005. During the benchmark at WKROUND (ICES, 2013), it was agreed that recent landings are likely to be more accurate from 2006 due to, i) less restrictive TACs, ii) the introduction of buyers and sellers legislation in the UK and Ireland and iii) the offshore gillnet fishery for anglerfish historically conducted by Spanish flagged vessels and thought to under-report landings, being much reduced. Anecdotal reports from fisheries officers and catch sampling staff suggest that towards the end of 2016 and into 2017 the high abundance of anglerfish on the grounds, and the restrictive quota were leading to an increase in suspected misreporting, discarding and black landings. There was no new information in 2021 to suggest that these suspected practices continued into 2018, 2019 or 2020, and the lower quota uptake during these years may indicate that the incentives for this behaviour are no longer prevalent. During the period 2005-2010, landings data were not provided to the Working Group by some of the major nations exploiting the fishery; however, the recent data call for the WKAngler benchmark (2018) has meant that WG estimates of subarea 6 and 4 landings have now been calculated for this period.

## Discards

Prior to the recent WKAngler benchmark (2018) discard estimates have only been available within InterCatch since 2012. Following the WKAngler data call discard information are now available for some fleets since 2002; however, discard information from UK (Scotland) is not available before 2008. The discard estimates that are available from other nations for the 20022007 period are substantially higher than the later UK (Scottish) rates. Given that these (nonScottish) fleets represent proportionally less of the landings, the discards pre-2008 are considered to be non-representative of the overall fishery (WKAngler 2018).

The breakdown of landings and discards by main gear group and area for 2020 and 2021 is given in Table 4.6. Landings and discards over time are shown in Figure 4.13. Discard data indicate that discarding in this fishery is relatively low due to high market value and no MCRS. Overall discarding was $2.4 \%$ of total catch in 2021. Demersal TR2 trawlers had the highest discard rate due to more restrictive quota share, $10.49 \%$ in 2021 , similar to the value for 2020 ( $12.7 \%$ ), but a substantial reduction from 2017 (20.9\%) and 2016 (43.9\%). In comparison TR1 trawlers, gillnets and miscellaneous gear types typically tend to have much lower discard rates (<2\%).

Figures $4.3(\mathrm{a}-\mathrm{c})$ show the percentage of landed weight by fleet, country and area. Length-frequency samples for catch in 2021 were submitted by Belgium, Denmark, France, Germany, Ireland, Norway, UK (England \& Wales) and the UK (Scotland). There was good coverage of both the demersal TR1 and TR2 fleets in Subarea 4 and Division 6.a. There were no samples from UKflagged gillnet vessels (operating in Subarea 4) which alone accounted for approximately $11 \%$ of all landings (Figure 4.3a).

## Length compositions

There is now a time-series of commercial catch-at-length data for 2002-2021 (Figure 4.4). The spread of lengths in the landings distributions are wider during the period 2012-2014 after which the distributions are steeper and unimodal. In 2015 the strong 2013 cohort entered the fishery producing a markedly different catch composition of lengths with the bulk of landings being between 30 and 50 cm in length with steep tails either side. Discard rates are lower from 2015 onwards however the landings of $<30 \mathrm{~cm}$ fish were also lower, suggesting this reduction could be a combination of catch composition and the increase in quota availability. The distribution of
lengths in the landings in 2020 and 2021 have narrowed when compared to 2018 and 2019 with low numbers of landings of fish $>75 \mathrm{~cm}$ notable in 2021.

## Biological Data

An anglerfish ageing exchange was held in 2011 to investigate the possibility of the collation of an international landings-at-age dataset of hard structure age readings, however little agreement was found between methods or readers. This was acknowledged in the findings of the WKROUND report on current assessment and issues with data and assessment of this stock (ICES, 2013). Further to this, discussions at WKAngler established that few countries are actively reading anglerfish hard structures, although they continue to be collected, processed and stored. It is unlikely that any developments in regards to an agreed reading criterion will be made in the near future.

## Research vessel surveys

The SIAMISS (Scottish Irish Anglerfish Megrim Industry Science Survey) is a dedicated anglerfish survey. It covers much of the known distribution of the northern shelf anglerfish (ICES divisions $4 \mathrm{a}, 6 \mathrm{a}$ and 6 b ), with the exception of the central and southern parts of Subarea 4 and the Skagerrak and Kattegat (Division 3a). The survey area has been stratified based on knowledge from fishermen with sampling effort within each stratum allocated roughly according to its expected biomass. Given the large spatial coverage of the survey, it is typically carried out by multiple vessels including commercial fishing vessels and both Irish and Scottish research vessels using a standard gear. Abundance and biomass estimates are worked up on the basis of swept-area and account for herding by the trawl doors and sweeps, ii) escapes under the footrope and iii) anglerfish abundance and biomass in the southern part of Division 6a were not covered in 2005, 2008 and 2010. Further details regarding the survey design and work up can be found in the stock annex and working document for 2021 (see Barreto et al., 2021).

The survey began in 2005 and is carried out on an annual basis (usually in spring, but sometimes in November). In 2020, however, the Scottish component of the SIAMISS survey (covering the northern North Sea, the north of divisions $6 a$ and $6 b$ ) was cancelled due to the COVID-19 pandemic. While the Irish part of the survey did go-ahead (covering the southern part of Division 6 a), historical densities and stock trends suggest that extrapolation of this component of the survey to the wider stock area would be inappropriate. Therefore, there is no abundance/biomass estimate from SIAMISS for 2020. In 2022 the anglerfish multi-vessel survey took place from the $12^{\text {th }}$ to $27^{\text {th }}$ of April and involved two vessels: FRV Scotia - surveying Division 4a and Division 6a North of $58^{\circ} \mathrm{N}$, and the Irish Marine Institute research vessel FRV Celtic Explorer, surveying Division 6a South of $58^{\circ} \mathrm{N}$. One haul with the duration of 60 minutes was made at each sampling station ( $\mathrm{n}=138$ ). Due to a mechanical fault with FRV Scotia the SIAMISS survey did not include Division 6b (Rockall) in 2022.

Figures 4.14 and 4.15 show the 2022 survey haul locations and mean numbers and weight per $\mathrm{km}^{2}$ caught at these locations. Larger numbers of anglerfish were caught along the shelf-edge below $58^{\circ} \mathrm{N}$ with large weights of fish being caught at the same locations. In previous years larger weights of fish have also been caught at Rockall, indicating that the fish at Rockall are larger than those caught on the shelf-edge. In 2022 there was no survey at Rockall so weights for this year are unknown.

Estimated total population numbers and biomass at length by area from the most recent survey in 2022 are shown in Figure 4.7 which show a much higher proportion of large fish in division $4 a$ than in division $6 a$. In terms of numbers, area $4 a$ has by far the highest value, when compared with areas 6 a . Comparison of numbers-at-length and weight-at-length over time for all areas
combined show a slight decrease in numbers but a slight increase in biomass compared to 2021. (Figures 4.8).

A time-series of total biomass is given in Table 4.7 and Figure 4.5. The total biomass estimate for the Northern Shelf in 2022, the most recent survey year was 55423 t an increase of $14 \%$ compared to 2021. A large proportion of total population numbers consisted of individuals $<40 \mathrm{~cm}$ in 2022, suggesting reasonably strong recruitment.

The breakdown of total numbers and biomass by area (Table 4.8 and Figure 4.6) shows that Division 6 b has lower estimated population numbers with less variability over time than in either division 6a or 4a. Division 4a consistently has the highest total biomass of the three areas and shows similar temporal trends to Division 6a.

Estimates of the ratio of survey biomass between subareas 4 and 6 have fluctuated around 1:1, (time-series average of $47.5 \%$ in Subarea 4, Table 4.7). The proportion of biomass in Subarea 4 had been steadily increasing since 2013; however, 2017 saw a slight decrease followed by a marked decline in 2018 to a time-series low of $37 \%$ (Figure 4.9). 2022 has seen an increase in the proportion of biomass in Subarea 4 moving back towards a $1: 1$ split.

Additional survey indices were developed during the WKAngler 2018 benchmark after revisiting the anglerfish abundance of several surveys within the stock area (ICES, 2018). Mean weight per hour for both the SCW-IBTS Q1 and Q4 surveys declined in 2018 following time-series highs in 2017 and 2016 respectively (Figure 4.12) which reflects the SIAMISS-Q2 biomass trend (Figure 4.6). The Rockall index (Figure 4.12) has shown an increasing trend since 2005 with a significant peak in 2012 followed by a short period of decline before a continuation of the increasing trajectory from 2016 to 2018, a slight increase again in 2019 and a slight decline in 2020. Although the SIAMISS-Q2 biomass time-series for Division 6. b shows less year to year fluctuation than the Rockall index the increasing trend and magnitude of change for the 2005-2021 period are very similar. In Subarea 4 the NS-IBTS-Q1 and Q3 indices showed declining mean weights per hour until 2020, with a slight increase in 2021 (Figure 4.13).

### 2.3 Historical stock development

There has been no analytic assessment of Northern Shelf anglerfish since 2003, due to a combination of unreliable commercial data, landings misreporting, uncertain effort data and poor catchability of anglerfish in traditional research surveys. The Scottish Irish anglerfish and megrim industry science survey (SIAMISS-Q2) initiated by Marine Scotland Science in 2005, along with official logbook data and tally-book data schemes have addressed some of these issues, providing valuable information to fishery managers as well as minimum absolute abundance and biomass estimates annually. Between 2012-2021 assessment followed the ICES RGLIFE datalimited stock (DLS) 3.2.0 method of survey based indicative trends (ICES, 2012). Beginning in 2022 the assessment followed the rfb assessment method proposed by WKLIFE (ICES, 2021).

At the benchmark in 2018, it was agreed to use SIAMISS-Q2 survey as an indicator of historical stock development. During the first half of the time-series, the biomass index for the stock fluctuated around 40 kt . Between 2011 and 2017, the total biomass increased significantly (more than doubled) due to very strong recruitment which is first observed in the survey length frequency data in 2013 and can continue to be clearly tracked through these data until 2017. Between 2017 and 2019, the estimated total stock biomass decreased by more than $30 \%$ before increasing slightly in 2022.

Figure 4.10 and Table 4.9 shows mean standardised harvest rate (calculated as catch/survey index) by both weight and number of individuals. Whilst there are no reference levels to relate these harvest rates to, trends can still be useful. In terms of biomass, the harvest rate has shown
an increasing trend since 2015 and in 2021 was estimated to be one of the highest values of the time-series. The harvest rate in number has shown a more gradual increase over this time. The marked fall in harvest rate by number from 2013-2014 is likely due to the influx of the substantial 2013 year class (i.e. large increase in survey numbers) and not a change in fishing behaviour. It may be more appropriate to use a harvest rate which is measured over a given length range of commercially exploitable fish.

### 2.4 Application of the rfb advice rule

Beginning in 2022 advice for this stock is given following the rfb assessment method (ICES, 2022) calculated as the previous year's advice multiplied by a) the index ratio (r) (the ratio of the mean of the last two index values (index A) and the mean of the three preceding values (index $B)$ ); b) a ratio of observed mean length in the catch relative to the target mean length (f); c) a biomass safeguard (b) calculated as either the ratio of the last index value and the index trigger value or 1 where this is lower; and d) a precautionary multiplier ( m ) based on life history which is fixed over time.

Rfb values are calculated as follows (ICES, 2021):

| Component | Definition | Description and use |
| :---: | :---: | :---: |
| $r$ | $\frac{\sum_{i=y-2}^{y-1}\left(I_{i} / 2\right)}{\sum_{i=y-5}^{y-3}\left(I_{i} / 3\right)}$ | The rate of change in the biomass index ( $I$ ), based on the average of the two most recent years of data ( $y-2$ to $y-1$ ) relative to the average of the three years prior to the most recent two $(y-3$ to $y-5$ ), and termed the " 2 over 3 " rule. |
| $f$ | $\frac{\bar{L}_{y-1}}{L_{F=M}}$ | The ratio of the mean length $\left(\bar{L}_{\mathrm{y}-1}\right)$ in the observed catch that is above the length of first capture relative to the target reference length (mean length/target reference length). The target reference length is $L_{F=M}=0.75 L_{c}+0.25 L_{\infty}$, where. $L_{c}$ is defined as length at $50 \%$ of modal abundance (ICES, 2018b). |
| $b$ | $\min \left\{1, \frac{I_{y-1}}{I_{\text {trigger }}}\right\}$ | Biomass safeguard. Adjustment to reduce catch when the most recent index data $I_{y-1}$ is less than $I_{\text {trigger }}=1.4 I_{\text {loss }}$ such that $b$ is set equal to $I_{y-1} / I_{\text {trigger }}$. When the most recent index data $I_{y-1}$ is greater than $I_{\text {trigger }}, b$ is set equal to $1 . I_{\text {loss }}$ is generally defined as the lowest observed index value for that stock. |
| $m$ | [0,1] | Multiplier applied to the harvest control rule to maintain the probability of the biomass declining below $B_{\text {lim }}$ to less than $5 \%$. May range from 0 to 1.0. |
| Stability clause | $\min \left\{\max \left(0.7 C_{y}, C_{y+1}\right), 1.2 C_{y}\right\}$ | Limits the amount the advised catch can change upwards or downwards between years. The recommended values are $+20 \%$ and $-30 \%$; i.e. the catch would be limited to a $20 \%$ increase or a $30 \%$ decrease relative to the previous year's advised catch. The stability clause does not apply when $b<1$. |

Due to the lack of SIAMISS survey data in 2020, the procedure for calculating the index ratio could not be followed exactly. Instead, the rfb framework for category 3 stocks is applied with the 2020 index value treated as missing. This adds a degree of uncertainty as without 2020 data index B has been calculated as the mean of the previous two years (2018 and 2019) instead of three years. Consequently, the 3 over 2 rule was replaced with a 2 over 2 rule in 2022. Furthermore, in 2022 the SIAMISS-Q2 survey did not include division 6.b (Rockall) which in 2021 accounted for $7.9 \%$ of landings. The survey estimate for 2021 was used in place of the missing 2022 estimate for division 6.b. This adds a further degree of uncertainty to the assessment, but does not affect the advice as the reduction is capped.

The advice in 2022 is lower than in 2021 due to a decrease in the index ratio (Figure 4.5), an $\mathrm{F}_{\text {MSY }}$ Proxy value lower than 1 (Figure 4.11) and a precautionary multiplier of 0.95 . The mean catch length of fish in this stock has consistently been below the proxy length since at least 2011 (Figure 4.11). Although biomass has declined since 2017 it has remained above the index trigger value since 2014 (Figure 4.5). The stability clause was considered and applied to limit the reduction in catch advice to $30 \%$.

| Division 3a46 Anglerfish stock |  |
| :--- | ---: |
| Previous catch advice Ay (advised catch for 2022) | 14116 tonnes |
| Stock biomass trend | 53222 tonnes |
| Index A (2021, 2022) | 68118 tonnes |
| Index B (2018, 2019; 2020 not available) | 0.78 |
| r: stock biomass trend (index ratio A/B) |  |

Fishing pressure proxy

| Mean catch length (Lmean = L2021) | 53 cm |
| :--- | ---: |
| MSY proxy length (LF $=\mathrm{m})$ | 61 cm |
| f: Fishing pressure proxy relative to MSY proxy (L2021/LF=m) | 0.87 |
| Biomass safeguard | 55423 |
| Last index value (I2021) | 46554 |
| Index trigger value (Itrigger $=$ Iloss $\times 1.4)$ | 1 |

Precautionary multiplier to maintain biomass above Blim with 95\% probability

| $\mathrm{m}:$ multiplier (generic multiplier based on life history) | 0.95 |
| :--- | :--- |


| RFB calculation** | 8893 tonnes |  |
| :--- | ---: | ---: |
| Stability clause (+20\%/-30\% compared to Ay, only applied if $b=1)$ | Applied | 0.70 |
| Discard rate | $1.2 \%$ |  |
| Catch advice for 2023 (Ay $\times$ stability clause) | 9881 tonnes |  |
| Projected landings corresponding to advice*** | 9610 tonnes |  |
| $\%$ advice change ${ }^{*}$ | $-30 \%$ |  |

[^1]
### 2.5 Biological reference points

Biological reference are calculated as follows:

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 46554 | Biomass index trigger value ( $I_{\text {trigger }}$ ), defined as $I_{\text {trig- }}$ ger $=I_{\text {loss }} \times 1.4$, where $I_{\text {loss }}$ is the lowest observed historical biomass index value (from 2011) from 2005-2022. Value in tonnes. | $\begin{gathered} \text { ICES } \\ (2022 a, \\ 2022 b) \end{gathered}$ |
|  | FMSY proxy | 0.87 | $L_{\text {mean }} / L_{F}=M$; Mean catch length divided by MSY proxy reference length ( $L_{F}=M=61 \mathrm{~cm}$ ). | ICES (2022a, $2022 b)$ |
| Precautionary approach | Blim | Not defined |  |  |
|  | $\mathrm{B}_{\mathrm{pa}}$ | Not defined |  |  |
|  | $\mathrm{F}_{\text {lim }}$ | Not defined |  |  |
|  | $\mathrm{F}_{\mathrm{pa}}$ | Not defined |  |  |
| Management plan | SSB ${ }_{\text {mgt }}$ | Not applicable |  |  |
|  | $F_{\text {mgt }}$ | Not applicable |  |  |

One suggested method for future assessment is a Nephrops-like harvest-ratio approach which creates a catch-options table based on a range of harvest ratios. However, to date no MSY reference points have been determined for Northern shelf anglerfish despite further exploration (Holah, H., 2017). Limited data, dome-shaped selectivity and uncertain life-history parameters continue to be inhibiting factors. Previous attempts to determine suitable harvesting rates, based on a yield-per-recruit analysis, estimated $\mathrm{Fmax}^{\max }$ to be 0.19 (ICES, 2004). The southern Celtic Sea/Bay of Biscay stock has recently been benchmarked and an FMSY of 0.28 was adopted (ICES, 2018a). In the case of Nephrops the technical basis for MSY Btrigger is the bias-adjusted lowest observed UWTV survey estimate of abundance, however for anglerfish, whilst abundances from SIAMISS-Q2 were initially intended to be an absolute measure of abundance they are now considered to be only a relative index so this may not be appropriate.

### 2.6 Management plans

ICES is aware of the multiannual management plan (MAP) which has been adopted by the EU for this stock (EU, 2019) and which ICES considers to be precautionary. There is no agreed shared management plan with the UK for this stock, and ICES provides advice according to ICES precautionary approach.

### 2.7 Uncertainties and bias in assessment and forecast

The WGCSE has previously attempted assessments of the anglerfish stock(s) within its remit using a number of different approaches. As yet none have proved entirely satisfactory. The catch-at-length analysis used in previous years appears to have addressed a number of the suspected problems with the data due to the rapid development of the fishery, and has also provided a satisfactory fit to the catch-at-length distribution data. However, since 2003, the WG has been
unable to present an analytic assessment due to the lack of reliable fishery and insufficient survey information, and in addition it is not known to what extent the dynamic pool assumptions of the traditional assessment model are valid for anglerfish. A catch-at-age model was presented to two benchmark working groups (WKFLAT 2012 and WKROUND 2013) but was not accepted due to concerns over age reading. The SPiCT and ASPIC surplus production models were explored at the WKAngler benchmark (2018) and whilst the models converged, the models were unstable and the uncertainty was large. This is most likely due to the lack of contrast in the catch data.

## Commercial data

For a number of years the WG has expressed concerns over the quality of the commercial catch-at-length data because of:

- Accuracy of landings statistics due to species and area misreporting (historically an issue between 1998-2005 and anecdotally again in 2016).
- Lack of information on total catch and catch composition of gillnetters operating on the continental slope to the northwest of the British Isles (See the stock annex for further details of this fishery).


## Survey data

There are still several factors which make the survey estimates likely to be underestimates or minimum estimates. Firstly, although experiments have been carried out to estimate escapes from under the footrope, and a model applied to account for this component of catchability, the estimates of smaller anglerfish still look to be underestimated (Figure 4.7). This could be due to either a net selectivity issue, or an availability [to the trawl] issue, as it is known that younger fish occur in shallower water (Hislop et al., 2001), or both. Secondly, the area considered is not complete, as the survey does not cover some of Division 4.a and none of 4.b or 4.c. However, numbers are thought to be low in these areas.

## Biological information

Knowledge of the biology of anglerfish has improved, with some basic biological parameters suitable for use in future assessments, such as mean weight-at-length in the stock, now available from the industry-science survey data. Difficulties still remain in finding mature females. A further discussion of the biology can be found in the stock annex.

Life-history parameters of the anglerfish species Lophius piscatorius and Lophius Budegassa in the Northeast Atlantic were reviewed at the WKAngler benchmark (2018) with appropriate ranges of natural mortality ( M ) discussed and new approaches to estimating age from mixture modelling of length distributions presented (see WKAngler 2018 report for further details).

## Stock structure

Currently, anglerfish on the Northern Shelf are split into Subarea 6 (including 5.b (EC), 7 and 14) and the North Sea (and 2.a (EC)) for management purposes. However, genetic studies have found no evidence of separate stocks over these two regions (including Rockall) and particletracking studies have indicated interchange of larvae between the two areas (Hislop et al., 2001). So, at previous WGs, assessments have been made for the whole Northern Shelf area combined. In fact, both microsatellite DNA analysis (O'Sullivan et al., 2005) and particle tracking studies carried out as part of EC 98/096 (Anon, 2001) also suggested that anglerfish from further south (Subarea 7) could also be part of the same stock.

At present, the stock is assessed for the two anglerfish species L. piscatorius and L. budegassa combined despite differing life-history characteristics and overlap in spatial distribution. This has been the case due to the black anglerfish (L. Budegassa) proportionally representing only around $10 \%$ of the estimated stock biomass from the SIAMISS-Q2 survey and that the Scottish fleet land the two species for sale combined as "monkfish". Given that the proportion of black anglerfish has been as high as $28 \%$ in Division 6.a and that the Scottish market sampling programme records to species level, a splitting out of black anglerfish in this stock may be a consideration for a future benchmark.

### 2.8 Recommendations for next Benchmark

This stock was last benchmarked in February 2018 at WKAngler. The recommendations to be carried forward following WKAngler are the following tasks:

- Investigate length-based stock assessment using, for example, the SS3 approach applied to southern anglerfish stocks.
- Investigate growth models appropriate for anglerfish subareas 4 and 6.
- Investigate an age-aggregated production/depletion model.
- Determine the best way to incorporate Lophius budegassa into assessment and advice.

The WKAngler data call led to the compilation of commercial sampling data (length, age, weight) previously held internationally, to construct a historical catch-at-length dataset for 2002 to present. At this stage, the focus is currently to utilise this more complete dataset to develop a suitable assessment model for this stock.

### 2.9 Management considerations

Up to and including 2011, ICES provided qualitative advice regarding the future exploitation of 'data-limited' stocks where there was either limited knowledge of their biology or a lack of data on their exploitation. However, in response to a strong interest from advice recipients to base advice on the information available, ICES developed the data-limited stocks (DLS) approach framework, for which anglerfish is a category 3 data-limited stock. This requires considering the application of an uncertainty cap and/or precautionary buffer to a survey adjusted status quo catch at each annual advice draft.

The two TACs in this area do not match the stock unit. One TAC area covers Subarea 4 and Division 2.a (EC); the second covers Division 5.b (EC) and subareas 6, 12, and 14. There is no TAC for Division 3.a and landings from this area have increased significantly in recent years. As a result of this mismatch, there is a potential for catches to exceed advice. There is no TAC for the Norwegian fishery in Subarea 4.

The TACs in subareas 4 (including Norwegian waters) and 6 until 2010 were split $67: 33 \%$, since 2011 they have been split $64: 36 \%$. In $2018,10 \%$ of the TAC for 4 and 2.a could be taken from Division 5.b, or subareas 6, 7 and 9. Over the survey time-series, the stock has been fairly evenly distributed between 4:6, the split has fluctuated around 50:50 ( $47 \%$ on average) (Table 4.7 and Figure 4.9) however in 2018 there was a significant decrease to $38 \%$ increasing to $40 \%$ in 2019 and $47 \%$ in 2021 . Note that the North Sea is only partially surveyed: however, the area covered does encompass most of the distribution of anglerfish.

Ideally, the management of the fishery should be based on a specific plan, or harvest control rule, after an evaluation of various stakeholder-led suggestions of alternative options. This still needs
to be pursued in consultation with stakeholders such as the North Western Waters Advisory Council.

### 2.10 References

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Table 4.1. ICES advice and actual management applicable for Northern Shelf anglerfish for 2003 onwards.

| YEAR | Catch corresponding <br> to advice | BASIS | WEST OF SCOTLAND (Sub- <br> area 6$)$ | NORTH SEA (Subarea 4) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2003 | TAC |  |  |  |

All values raised to nearest tonne.
${ }^{\wedge}$ Landings advice

1) Advice for Division 3.a, Subarea 4 and Subarea 6.a combined.
2) Advice for Division 3.a, Subarea 4 and Subarea 6 combined.
3) Advice for Division 2.a, Division 3.a, Subarea 4 and Subarea 6 combined.
4) pplies to $5 . b$ (EC), 6, 7 and 14.
5) TAC applies to 2.a \& 4 (EC).
(7) Landings including raised discards.

Although there is no minimum landing size for this species, there is an EU minimum weight of 500 g for marketing purposes (EC Regulation 2406/96).

An additional quota of 1500 t was also available for EU vessels fishing in the Norwegian zone of Subarea 4 in 2011-2018 which was increased to 1700 t in 2018.

Table 4.2. Anglerfish on the Northern Shelf (3.a, $4 \& 6$ ). Total official landings by area (tonnes).

|  | 3.2 | 4.a | 4.b | 4.c | $6 . a$ | 6.b | 4 | 6 | Total (3.A, 4,6) | WG Landings | WG Discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1973 | 140 | 2085 | 575 | 41 | 9221 | 127 | 2701 | 9348 | 12189 | - | - |
| 1974 | 202 | 2737 | 1171 | 39 | 3217 | 435 | 3947 | 3652 | 7801 | - | - |
| 1975 | 291 | 2887 | 1864 | 59 | 3122 | 76 | 4810 | 3198 | 8299 | - | - |
| 1976 | 641 | 3624 | 1252 | 49 | 3383 | 72 | 4925 | 3455 | 9021 | - | - |
| 1977 | 643 | 3264 | 1278 | 54 | 3457 | 78 | 4596 | 3535 | 8774 | - | - |
| 1978 | 509 | 3111 | 1260 | 72 | 3117 | 103 | 4443 | 3220 | 8172 | - | - |
| 1979 | 687 | 2972 | 1578 | 112 | 2745 | 29 | 4662 | 2774 | 8123 | - | - |
| 1980 | 652 | 3450 | 1374 | 175 | 2634 | 200 | 4999 | 2834 | 8485 | - | - |
| 1981 | 549 | 2472 | 752 | 132 | 1387 | 331 | 3356 | 1718 | 5623 | - | - |
| 1982 | 529 | 2214 | 654 | 99 | 3154 | 454 | 2967 | 3608 | 7104 | - | - |
| 1983 | 506 | 2465 | 1540 | 181 | 3417 | 433 | 4186 | 3850 | 8542 | - | - |
| 1984 | 568 | 3874 | 1803 | 188 | 3935 | 707 | 5865 | 4642 | 11075 | - | - |
| 1985 | 578 | 4569 | 1798 | 77 | 4043 | 1013 | 6444 | 5056 | 12078 | - | - |
| 1986 | 524 | 5594 | 1762 | 47 | 3090 | 1326 | 7403 | 4416 | 12343 | - | - |
| 1987 | 589 | 7705 | 1768 | 66 | 3955 | 1294 | 9539 | 5249 | 15377 | - | - |
| 1988 | 347 | 7737 | 2061 | 95 | 6003 | 1730 | 9893 | 7733 | 17973 | - | - |
| 1989 | 334 | 7868 | 2121 | 86 | 5729 | 313 | 10075 | 6042 | 16451 | - | - |
| 1990 | 570 | 8387 | 2177 | 34 | 5615 | 822 | 10598 | 6437 | 17605 | - | - |
| 1991 | 595 | 9235 | 2522 | 26 | 5061 | 923 | 11790 | 5984 | 18369 | 17441 | - |
| 1992 | 938 | 10209 | 3053 | 39 | 5479 | 1089 | 13301 | 6568 | 20807 | 21872 | - |
| 1993 | 843 | 12309 | 3143 | 66 | 5553 | 681 | 15519 | 6234 | 22596 | 23971 | - |
| 1994 | 811 | 14505 | 3445 | 210 | 5273 | 909 | 18162 | 6182 | 25155 | 25057 | - |
| 1995 | 823 | 17891 | 2627 | 402 | 6354 | 958 | 20920 | 7312 | 29055 | 28913 | - |
| 1996 | 702 | 25176 | 1847 | 304 | 6408 | 602 | 27327 | 7010 | 35039 | 35100 | - |


|  | 3.a | $4 . \mathrm{a}$ | 4.b | 4.c | 6.a | 6.b | 4 | 6 | Total (3.A, 4,6) | WG Landings | WG Discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 776 | 23425 | 2172 | 160 | 5330 | 990 | 25757 | 6320 | 32853 | 32728 | - |
| 1998 | 626 | 16859 | 2088 | 78 | 4506 | 1313 | 19026 | 5819 | 25471 | 25293 | - |
| 1999 | 660 | 13344 | 1517 | 24 | 4284 | 1401 | 14885 | 5685 | 21230 | 21854 | - |
| 2000 | 602 | 12338 | 1617 | 31 | 3311 | 1074 | 13986 | 4385 | 18973 | 19682 | - |
| 2001 | 621 | 12861 | 1832 | 21 | 2660 | 1309 | 14714 | 3969 | 19304 | 19157 | - |
| 2002 | 667 | 11048 | 1244 | 21 | 2280 | 718 | 12313 | 2998 | 15978 | 15067 | - |
| 2003 | 478 | 8523 | 847 | 20 | 2493 | 643 | 9390 | 3136 | 13004 | 12008 | - |
| 2004 | 519 | 8987 | 851 | 15 | 2453 | 671 | 9853 | 3124 | 13496 | 11976 | - |
| 2005 | 458 | 8424 | 688 | 5 | 3019 | 958 | 9117 | 3982 | 13557 | 13728 | - |
| 2006 | 426 | 10340 | 683 | 3 | 2785 | 915 | 11026 | 3700 | 15152 | 13292 | - |
| 2007 | 433 | 10632 | 749 | 4 | 3353 | 1261 | 11384 | 4613 | 16430 | 14564 | 490 |
| 2008 | 486 | 11038 | 769 | 5 | 3373 | 1246 | 11813 | 4619 | 16918 | 15878 | 903 |
| 2009 | 478 | 10067 | 651 | 8 | 2984 | 1820 | 10726 | 4804 | 16008 | 15372 | 38 |
| 2010 | 433 | 8190 | 615 | 11 | 3040 | 1606 | 8815 | 4645 | 13895 | 12136 | 69 |
| 2011 | 405 | 7760 | 764 | 8 | 2871 | 1871 | 8532 | 4742 | 13679 | 12902 | 95 |
| 2012 | 423 | 6459 | 714 | 4 | 2835 | 1831 | 7177 | 4666 | 12266 | 11143 | 590 |
| 2013 | 407 | 6393 | 546 | 5 | 2667 | 2123 | 6944 | 4790 | 12141 | 11375 | 687 |
| 2014 | 440 | 7633 | 820 | 27 | 2610 | 1754 | 8481 | 4365 | 13286 | 14406 | 448 |
| 2015 | 478 | 9690 | 985 | 16 | 3290 | 1723 | 10691 | 5013 | 16182 | 15663 | 395 |
| 2016 | 586 | 11680 | 1196 | 11 | 4638 | 1423 | 12887 | 6060 | 19533 | 19412 | 981 |
| 2017 | 742 | 13620 | 1107 | 7 | 5024 | 1504 | 14733 | 6528 | 22023 | 21719 | 756 |
| 2018 | 914 | 13438 | 823 | 11 | 4369 | 1932 | 14274 | 6303 | 21487 | 21572 | 326 |
| 2019 | 1029 | 11155 | 1303 | 28 | 5030 | 2647 | 12486 | 7677 | 21192 | 21- | 513 |
| 2020* | 886 | 10674 | 1066 | 24 | 4749 | 1685 | 11763 | 6434 | 19064 | 19072 | 316 |
| 2021* | 912 | 11959 | 1184 |  | 19 | 4455 | 1537 | 13162 | 5992 | 20066 | 20143 |

*Preliminary.

Table 4.3. Anglerfish in Subarea 6. Nominal landings ( $\mathbf{t}$ ) as officially reported to ICES.
$\underline{\text { Division 6.a (West of Scotland) }}$

|  | ন্ন | ন্ন | $\stackrel{\text { ® }}{-}$ | 茳 |  | ® | $\stackrel{\text { N}}{-}$ | ®̊ | ু | O- ㅁN | ત্ণ | 꿍 | ÒN | ষ্ণ | NiN | Ò O | 웅 | ÒN | OiO | 익 | $\underset{\sim}{\underset{\sim}{7}}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{\underset{N}{N}}$ | $\underset{\sim}{\underset{N}{N}}$ | $\stackrel{n}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{N}{n} \end{aligned}$ | $\stackrel{\underset{\sim}{N}}{ }$ | $\stackrel{\infty}{\underset{\sim}{7}}$ | İ물 | $\begin{aligned} & \text { * } \\ & \text { N్N } \end{aligned}$ | $\stackrel{*}{\underset{\sim}{\sim}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 | 2 | 9 | 6 | 5 | - | 5 | 2 | - | - | + | + | - | + | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Den- <br> mark | 1 | 3 | 4 | 5 | 10 | 4 | 1 | 2 | 1 | + | + | - | + | + | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 1 |
| Faroe Is. | - | - | - | - | - | - | - | - | - | - | - | - | - | 2 | 2 | 3 | 2 | 1 | 2 | 4 | 1 | - | - | - | + | 1 | - | - | - | - | - |
| France | $\begin{aligned} & 191 \\ & 0 \end{aligned}$ | $\begin{aligned} & 230 \\ & 8 \end{aligned}$ | $\begin{aligned} & 246 \\ & 7 \end{aligned}$ | $\begin{aligned} & 238 \\ & 2 \end{aligned}$ | 2648 | 2899 | 2058 | $\begin{aligned} & 163 \\ & 4 \end{aligned}$ | $\begin{aligned} & 181 \\ & 4 \end{aligned}$ | $\begin{aligned} & 113 \\ & 2 \end{aligned}$ | 943 | 739 | $\begin{aligned} & 121 \\ & 2 \end{aligned}$ | $\begin{aligned} & 119 \\ & 1 \end{aligned}$ | $\begin{aligned} & 139 \\ & 6 \end{aligned}$ | $\begin{aligned} & 131 \\ & 4 \end{aligned}$ | $\begin{aligned} & 176 \\ & 4 \end{aligned}$ | $\begin{aligned} & 174 \\ & 6 \end{aligned}$ | $\begin{aligned} & 151 \\ & 3 \end{aligned}$ | $\begin{aligned} & 120 \\ & 6 \end{aligned}$ | $116$ | $\begin{aligned} & 116 \\ & 6 \end{aligned}$ | $\begin{aligned} & 111 \\ & 4 \end{aligned}$ | $\begin{aligned} & 109 \\ & 8 \end{aligned}$ | $\begin{aligned} & 110 \\ & 7 \end{aligned}$ | $\begin{aligned} & 173 \\ & 4 \end{aligned}$ | $\begin{aligned} & 188 \\ & 2 \end{aligned}$ | $\begin{aligned} & 128 \\ & 7 \end{aligned}$ | $\begin{aligned} & 127 \\ & 6 \end{aligned}$ | $\begin{aligned} & 12 \\ & 81 \end{aligned}$ | 1322 |
| Germany | 1 | 2 | 60 | 67 | 77 | 35 | 72 | 137 | 50 | 39 | 11 | 3 | 27 | 39 | 39 | 1 | - | 54 | 79 | 79 | 59 | 63 | 48 | 85 | 63 | 81 | 79 | 127 | 94 | 8 | 35 |
| Ireland | 250 | 403 | 428 | 303 | 720 | 717 | 625 | 749 | 617 | 515 | 475 | 304 | 322 | 219 | 356 | 392 | 470 | 295 | 328 | 510 | 488 | 346 | 336 | 410 | 446 | 581 | 579 | 596 | 897 | $\begin{aligned} & 69 \\ & 8 \end{aligned}$ | 517 |
| NL | - | - | - | - | - | - | 27 | 1 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Norway | 6 | 14 | 8 | 6 | 4 | 4 | 1 | 3 | 1 | 3 | 2 | 1 | + | + | 1 | 1 | 1 | 2 | + | 2 | 1 | + | 1 | 1 | 1 | 1 | 1 | 2 | 2 | 1 | - |
| Spain | 7 | 11 | 8 | 1 | 37 | 33 | 63 | 86 | 53 | 82 | 70 | 101 | 196 | 110 | 83 | 76 | 3 | 174 | 185 | 197 | 138 | 69 | 123 | 54 | 30 | 178 | 173 | 218 | 298 | $\begin{aligned} & 23 \\ & 2 \end{aligned}$ | 251 |
| UK(E,W $\& N I)$ | 270 | 351 | 223 | 370 | 320 | 201 | 156 | 119 | 60 | 44 | 40 | 32 | 31 | 30 | 20 | 24 | 42 | 5 | 12 | 3 | - | 12 | 6 | - | - | - | - | - | - | - | - |
| UK(Scot .) | 261 3 | 238 5 | 234 6 | 213 3 | 2533 | 2515 | 2322 | $\begin{aligned} & 177 \\ & 3 \end{aligned}$ | 168 8 | $\begin{aligned} & 149 \\ & 6 \end{aligned}$ | $\begin{aligned} & 111 \\ & 9 \end{aligned}$ | $\begin{aligned} & 110 \\ & 0 \end{aligned}$ | 705 | 862 | $\begin{aligned} & 112 \\ & 7 \end{aligned}$ | 974 | $\begin{aligned} & 107 \\ & 1 \end{aligned}$ | $\begin{aligned} & 109 \\ & 6 \end{aligned}$ | 864 | $\begin{aligned} & 104 \\ & 0 \end{aligned}$ | - | $\begin{aligned} & 117 \\ & 9 \end{aligned}$ | $\begin{aligned} & 103 \\ & 8 \end{aligned}$ | - | - | - | - | - | - | - | - |
| UK (total) | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | $\begin{aligned} & 101 \\ & 6 \end{aligned}$ |  |  | 962 | $\begin{aligned} & 164 \\ & 3 \end{aligned}$ | $\begin{aligned} & 206 \\ & 2 \end{aligned}$ | $\begin{aligned} & 231 \\ & 1 \end{aligned}$ | $\begin{aligned} & 213 \\ & 9 \end{aligned}$ | $\begin{aligned} & 246 \\ & 3 \end{aligned}$ | $\begin{aligned} & 25 \\ & 29 \end{aligned}$ | 2330 |


|  | ন্ন | N్ | $\stackrel{๊}{\sigma}$ | $\underset{~}{\text { G }}$ | 늑 | இ. | 人 | ®̊ | ${ }^{\circ}$ | ిod | O్ત | O్N | ÒN | ষ্N | 告 | O O | 잉 | Oi | Oio | O- | $\stackrel{\underset{\sim}{1}}{ }$ | $\underset{\sim}{\sim}$ | $\stackrel{\sim}{N}$ | $\underset{\sim}{\underset{\sim}{A}}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\rightharpoonup}{\sim}$ | $\stackrel{\underset{\sim}{N}}{ }$ | $\stackrel{\infty}{\sim}$ | $\underset{\sim}{\circ}$ | $\begin{aligned} & \text { * } \\ & \text { त्N } \end{aligned}$ | $\stackrel{*}{\underset{\sim}{N}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total | 506 | 547 | 555 | 527 | 6354 | 6408 | 5330 | $\begin{aligned} & 450 \\ & 6 \end{aligned}$ | $\begin{aligned} & 428 \\ & 4 \end{aligned}$ | $\begin{aligned} & 331 \\ & 1 \end{aligned}$ | $\begin{aligned} & 266 \\ & 0 \end{aligned}$ | $\begin{aligned} & 228 \\ & 0 \end{aligned}$ | $\begin{aligned} & 249 \\ & 3 \end{aligned}$ | $\begin{aligned} & 245 \\ & 3 \end{aligned}$ | $\begin{aligned} & 302 \\ & 4 \end{aligned}$ | $\begin{aligned} & 278 \\ & 5 \end{aligned}$ | $\begin{aligned} & 335 \\ & 3 \end{aligned}$ | $\begin{aligned} & 337 \\ & 3 \end{aligned}$ | $\begin{aligned} & 298 \\ & 4 \end{aligned}$ | 3040 | 2871 | 2835 | 2667 | 261 | 329 | 463 | 502 | 436 | 503 | 47 | 4456 |
|  | 1 | 9 | 3 | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 8 | 4 | 9 | 0 | 49 |  |
| Unallo－ cated | 296 | $\begin{aligned} & 263 \\ & 8 \end{aligned}$ | $\begin{aligned} & 381 \\ & 6 \end{aligned}$ | $\begin{aligned} & 276 \\ & 6 \end{aligned}$ | 5112 | $\begin{aligned} & 1114 \\ & 8 \end{aligned}$ | 7506 | $\begin{aligned} & 523 \\ & 4 \end{aligned}$ | $\begin{aligned} & 379 \\ & 9 \end{aligned}$ | $\begin{aligned} & 311 \\ & 4 \end{aligned}$ | $\begin{aligned} & 206 \\ & 8 \end{aligned}$ | 187 | 2 | 16 | －8 | －74 | 145 | 332 | 190 | 56 | 62 | 91 | 115 | $\begin{aligned} & 159 \\ & 5 \end{aligned}$ | 68 | －58 | 12 | 42 | 290 | －6 | 4 |
| As used | 535 | 811 | 936 | 803 | 1146 | 1755 | 1283 | 974 | 808 | 642 | 472 | 246 | 249 | 246 | 301 | 271 | 349 | 370 | 317 | 309 | 293 | 292 | 278 | 420 | 335 | 458 | 503 | 441 | 532 | 47 | 4451 |
| by WG | 7 | 7 | 9 | 9 | 6 | 6 | 6 | 0 | 3 | 5 | 8 | 7 | 5 | 9 | 6 | 1 | 8 | 5 | 4 | 6 | 3 | 6 | 2 | 5 | 8 | 0 | 6 | 1 | 0 | 43 |  |

## ＊Preliminary．

## Table 4．3．Continued．Anglerfish in Subarea 6．Nominal landings（t）as officially reported to ICES．

Division 6．b（Rockall）／＂Preliminary．

|  | ন্ন | N্ন | 끅 | 茳 | ๗ু | ঃ | 측 | $\stackrel{\circ}{\circ}$ | 윽 | Oi | 무N | 우N | 잉 | 후N | 승 | O O N | 웅 | OiN | 웅 | Oì | $\underset{\sim}{7}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{\mathrm{N}}$ | $\underset{\sim}{J}$ | $\stackrel{\text { ñ }}{\text { N }}$ | $\begin{gathered} 0 \\ \underset{\sim}{\circ} \end{gathered}$ | $\stackrel{\text { Ni }}{\text { N }}$ | $\stackrel{\infty}{\underset{\sim}{N}}$ | $\stackrel{9}{\sim}$ | $\begin{gathered} \stackrel{*}{0} \\ \text { ON } \end{gathered}$ | $\stackrel{*}{\sim}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Is． | － | 2 | － | － | － | 15 | 4 | 2 | 2 | － | 1 | － | － | － | － | － | － | 1 | 4 | 8 | － | 5 | － | 1 | ＋ | ＋ | ＋ | － | － | － | － |
| France | － | － | 29 | － | － | － | 1 | 1 | － | 48 | 192 | 43 | 191 | 175 | 293 | 224 | 327 | 327 | 339 | 168 | 508 | 456 | 663 | 148 | 219 | － | － | － | 17 | 23 | 4 |
| Germany | － | － | 103 | 73 | 83 | 78 | 177 | 132 | 144 | 119 | 67 | 35 | 64 | 66 | 77 | 72 | 222 | 93 | 132 | 87 | 90 | 79 | 88 | 66 | 139 | 177 | 167 | 266 | 340 | 221 | 122 |
| Ireland | 272 | 417 | 96 | 135 | 133 | 90 | 139 | 130 | 75 | 81 | 134 | 51 | 26 | 13 | 35 | 53 | 70 | 76 | 91 | 107 | 108 | 235 | 237 | 162 | 156 | 160 | 214 | 282 | 365 | 202 | 167 |
| Norway | 18 | 10 | 17 | 24 | 14 | 11 | 4 | 6 | 5 | 11 | 5 | 3 | 6 | 5 | 4 | 6 | 7 | 5 | 9 | 12 | 7 | 5 | 9 | 3 | 6 | 11 | 4 | 1 | 1 | 2 | － |
| Portugal | － | － | － | 132 | 128 | － | 91 | 413 | 429 | 20 | 18 | 8 | 4 | 19 | 63 | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Russia | － | － | － | － | － | － | － | － | － | － | 1 | － | － | 2 | 4 | 1 | 1 | 35 | － | － | － | － | － | 1 | 2 | － | 2 | － | 1 | 3 | － |
| Spain | 333 | 263 | 178 | 214 | 296 | 196 | 171 | 252 | 291 | 149 | 327 | 128 | 59 | 43 | 34 | 36 | 12 | 85 | 57 | 32 | 29 | 36 | － | 27 | 119 | 56 | 118 | 43 | 60 | 32 | 31 |


|  | ন্ন | Nั | $\stackrel{m}{\sigma}$ | す | 능 | 윽 | 욱 | $\stackrel{\infty}{\circ}$ | 合 | O- 우 | O-N | N | Ò | Oì | in ì | 움 | ిస్N | Oi | Oì | 음 | $\underset{\sim}{\underset{\sim}{7}}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{\underset{\sim}{n}}$ | $\underset{\sim}{\underset{N}{N}}$ | $\stackrel{n}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{N}{1} \end{aligned}$ | $\stackrel{N}{\underset{N}{N}}$ | $\stackrel{\infty}{\underset{\sim}{N}}$ | $\stackrel{\square}{\sim}$ | ＊ | $\stackrel{*}{\text {－}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK（E，W\＆NI） | 99 | 173 | 76 | 50 | 105 | 144 | 247 | 188 | 111 | 272 | 197 | 133 | 133 | 54 | 93 | 46 | － | 1 | 48 | 15 | － | 120 | 395 | － | － | － | － | － | － | － | － |
| UK（Scot） | 201 | 224 | 182 | 281 | 199 | 68 | 156 | 189 | 344 | 374 | 367 | 317 | 160 | 294 | 355 | 477 | － | 624 | 1141 | 1177 | － | 895 | 732 | － | － | － | － | － | － | － | － |
| UK（total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | 622 | － | － | － | 1129 |  |  | 1347 | 1081 | 1018 | 999 | 1340 | 1862 | 1202 | 1209 |
| Total | 923 | 1089 | 681 | 909 | 958 | 602 | 990 | 1313 | 1401 | 1074 | 1309 | 718 | 643 | 671 | 958 | 915 | 1261 | 1246 | 1820 | 1606 | 1871 | 1831 | 2123 | 1754 | 1723 | 1423 | 1504 | 1946 | 2632 | 1685 | 1533 |
| Unallocated | － | － | － | －132 | －128 | － | －91 | －413 | －9 | 17 | －178 | 210 | 70 | 10 | 227 | 136 | 282 | 104 | －198 | 791 | 111 | 385 | 178 | 80 | 74 | 37 | 80 | 2 | 140 | －115 | －28 |
| As used by WG | 923 | 1089 | 681 | 777 | 830 | 602 | 899 | 900 | 1392 | 1091 | 1131 | 508 | 573 | 661 | 731 | 779 | 979 | 1142 | 2018 | 815 | 1760 | 1446 | 1945 | 1674 | 1649 | 1386 | 1424 | 1944 | 2492 | 1800 | 1561 |

## Table 4．3．Continued．Anglerfish in Subarea 6．Nominal landings（ $\mathbf{t}$ ）as officially reported to ICES．

## Subarea 6 （West of Scotland and Rockall）

${ }^{\wedge}$ indicates landings assigned to subarea 6 but not to a division．${ }^{*}$ Preliminary．

|  | ন্ন | Nু | $\stackrel{\text { ® }}{\underset{~}{2}}$ | ホ | №묵 | : | - | ®̊ | 稌 | O O | -্~ㅇ | 웅 | ÒN | O | ®in | O O | 웅 | OiN | Oio | 음 | $\underset{\sim}{\underset{\sim}{1}}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{\underset{\sim}{2}}$ | $\underset{\sim}{\underset{\sim}{I}}$ | $\stackrel{n}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{\sim}{n} \end{aligned}$ | Nì | $\stackrel{\infty}{\sim}$ | $\stackrel{\text { İ }}{ }$ | $\begin{aligned} & \text { * } \\ & \text { N్N } \end{aligned}$ | $\stackrel{\text { N }}{\text { N }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 | 2 | 9 | 6 | 5 | － | 5 | 2 | － | － | ＋ | ＋ | － | ＋ | － | － | － | － | － | － | － | － | － | － | － | － | －－ | － | － | － | － |
| Denmark | 1 | 3 | 4 | 5 | 10 | 4 | 1 | 2 | 1 | ＋ | ＋ | － | ＋ | ＋ | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | 1 |
| Faroe ls． | － | 2 | － | － | － | 15 | 4 | 2 | 2 | － | 1 | － | － | 2 | 2 | 3 | 2 | 2 | 6 | 12 | 1 | 5 | － | 1 | ＋ | 1 | ＋ | － | － | － | － |
| France | 191 | 230 | 2496 | 238 | 2648 | 2899 | 2059 | 1635 | 181 | 118 | 113 | 782 | 140 | 136 | 1689 | 153 | 209 | 207 | 185 | 137 | 167 | 162 | 177 | 124 | 132 | 173 | 188 | 128 | 1293 | 1 | 1326 |
| Germany | 1 | 2 | 163 | 140 | 160 | 113 | 249 | 269 | 194 | 158 | 78 | 38 | 91 | 105 | 116 | 73 | 222 | 146 | 211 | 166 | 149 | 142 | 136 | 151 | 201 | 258 | 246 | 394 | 434 | 2 | 157 |
| Ireland | 522 | 820 | 524 | 438 | 853 | 807 | 764 | 879 | 692 | 596 | 609 | 355 | 348 | 232 | 391 | 445 | 540 | 371 | 419 | 617 | 596 | 581 | 572 | 572 | 602 | 741 | 793 | 878 | 1262 | 8 | 684 |


|  | ন্ন | Nั | $\stackrel{\text { ® }}{-}$ | す | に | Һ | ন্ন | $\stackrel{\circ}{\underset{\sim}{\circ}}$ | 익 | Oi | Ö | N | 응 | ষ্N | 잉 | 웅 | 숭 | Oi | 웅 | 음 | $\underset{\sim}{7}$ | $\underset{\sim}{\sim}$ | $\stackrel{m}{N}$ | $\underset{\sim}{\underset{\sim}{J}}$ | $\stackrel{n}{\sim}$ | $\begin{aligned} & 0 \\ & \stackrel{\rightharpoonup}{N} \end{aligned}$ | $\stackrel{\underset{\sim}{i}}{ }$ | $\stackrel{\infty}{\sim}$ | $\stackrel{\overbrace{}}{\mathrm{N}}$ | $\begin{gathered} \stackrel{*}{\sim} \\ \text { DN } \end{gathered}$ | $\stackrel{*}{\text { N }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Nether- | - | - | - | - | - | - | 27 | 1 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - - | - |
| Norway | 24 | 24 | 25 | 30 | 18 | 15 | 5 | 9 | 6 | 14 | 7 | 4 | 6 | 5 | 5 | 7 | 8 | 7 | 9 | 14 | 7 | 6 | 10 | 4 | 8 | 12 | 5 | 4 | 2 | 1 | - |
| Portugal | - | - | - | 132 | 128 | - | 91 | 413 | 429 | 20 | 18 | 8 | 4 | 19 | 63 | - | - | - | - | - | - | - | - | - | - | - |  | - | - | - - | - |
| Russia | - | - | - | - | - | - | - | - | - | - | 1 | - | - | 2 | 4 | 1 | 1 | 35 | - | - | - | - | - | 1 | 2 | - | 2 | - | 1 | 3 | - |
| Spain | 340 | 274 | 186 | 215 | 333 | 229 | 234 | 338 | 344 | 231 | 397 | 229 | 255 | 153 | 117^ | 112 | 15 | 259 | 242 | 229 | 167 | 105 | 123 | 81 | 149 | 234 | 290 | 261 | 358 | 2 | 282 |
| UK(E,W\&NI | 369 | 524 | 299 | 420 | 425 | 345 | 403 | 307 | 171 | 316 | 237 | 165 | 164 | 84 | 113 | 70 | 188 | 6 | 60 | - | - | 132 | 401 | - | - | - | - | - | - | - - | - |
| UK(Scot) | 281 | 260 | 2528 | 241 | 2732 | 2583 | 2478 | 1962 | 203 | 187 | 148 | 141 | 865 | 115 | 1482 | 145 | 154 | 172 | 200 | - | - | 207 | 177 | - | - | - | - | - | - | - | - |
| UK (total) | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 223 | 214 | 220 | 217 | 231 | 272 | 308 | 331 | 347 | 4322 | 3 | 3539 |
| Total | 598 | 656 | 6234 | 618 | 7312 | 7010 | 6320 | 5819 | 568 | 438 | 396 | 299 | 313 | 312 | 3982 | 370 | 461 | 461 | 480 | 464 | 474 | 466 | 479 | 436 | 501 | 606 | 652 | 630 | 7652 | 6 | 5989 |
| Unallo- | 296 | 263 | 3816 | 263 | 4984 | 1114 | 7415 | 4821 | 379 | 313 | 189 | 22 | 68 | 6 | 235 | 209 | 137 | 228 | 388 | 733 | 49 | 294 | 63 | 151 | 5 | 94 | 68 | -53 | -160 | - | -23 |
| As used by | 628 | 920 | 1005 | 881 | 1229 | 1815 | 1373 | 1064 | 947 | 751 | 585 | 297 | 306 | 313 | 3747 | 349 | 447 | 484 | 519 | 391 | 469 | 437 | 472 | 588 | 500 | 596 | 646 | 635 | 7812 | 6 | 6012 |
| WG | 0 | 6 | 0 | 6 | 6 | 8 | 5 | 0 | 5 | 6 | 9 | 6 | 8 | 0 |  | 1 | 6 | 7 | 2 | 2 | 3 | 2 | 7 | 0 | 8 | 6 | 0 | 6 |  | 5 |  |

Table 4.4. Nominal landings ( t ) of Anglerfish in the North Sea, as officially reported to ICES.
Northern North Sea (4.a)

| BE | 2 | 9 | 3 | 3 | 2 | 8 | 4 | 1 | 5 | 12 | - | 8 | 1 | - | - | - | - | - | - | - | - | - | + | - | - | - | 1 | - | - | - | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DK | 124 | 1265 | 946 | 1157 | 732 | 1239 | 1155 | 1024 | 1128 | 1087 | 1289 | 1308 | 152 | 153 | 137 | 1311 | 961 | 1071 | 1134 | 114 | 841 | 821 | 854 | 801 | 962 | 1506 | 2002 | 1790 | 1669 | 1058 | 1140 |
|  | 5 |  |  |  |  |  |  |  |  |  |  |  | 3 | 8 | 9 |  |  |  |  | 3 |  |  |  |  |  |  |  |  |  |  |  |
| Faroes | 1 | - | 10 | 18 | 20 | - | 15 | 10 | 6 | - | 2 | - | 3 | 11 | 22 | 2 | - | - | 4 | - | - | - | - | - | - | - | - | - | 1 | - | - |
| FR | 124 | 151 | 69 | 28 | 18 | 7 | 7 | 3 | 18 | 8 | 9 | 8 | 8 | 8 | 4 | 7 | 13 | 13 | 20 | 23 | 20 | 14 | 15 | 27 | 26 | 35 | 91 | 141 | 185 | 124 | 108 |


|  | ন্ন | ুㅜㄱ | $\stackrel{\text { m }}{-}$ | す্ন | ฝু | 윽 | - | $\stackrel{\circ}{\mathrm{O}}$ | 名 | OㅇN | Ö | N్ N | Ò No | ষ্ণ | N్N | O O | 을 | O~N | Oi | 음 | $\underset{\sim}{7}$ | $\underset{\sim}{\sim}$ | $\stackrel{\sim}{N}$ | $\underset{\sim}{\underset{N}{N}}$ | $\stackrel{\text { ñ }}{\sim}$ | $\begin{aligned} & 0 \\ & \stackrel{\rightharpoonup}{N} \end{aligned}$ | ̇ㅗㅇ | $\stackrel{\sim}{\sim}$ | $\stackrel{\overbrace{}}{2}$ | $\stackrel{*}{\sim}$ | $\stackrel{*}{\sim}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DE | 71 | 68 | 100 | 84 | 613 | 292 | 601 | 873 | 454 | 182 | 95 | 95 | 65 | 20 | 84 | 173 | 186 | 344 | 216 | 124 | 46 | 265 | 274 | 321 | 286 | 208 | 523 | 462 | 547 | － | － |
| NL | 23 | 44 | 78 | 38 | 13 | 25 | 12 | － | 15 | 12 | 3 | 8 | 9 | 38 | 13 | 14 | 14 | 12 | 5 | 8 | 5 | 5 | － | 16 | － | 21 | 28 | 68 | 68 | 40 | 10 |
| No | 587 | 635 | 1224 | 1318 | 657 | 821 | 672 | 954 | 1219 | 1182 | 1212 | 928 | 769 | 999 | 880 | 1006 | 831 | 860 | 859 | 791 | 494 | 485 | 545 | 524 | 406 | 610 | 840 | 1230 | 1267 | 954 | 964 |
| ES | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | ＋ | － | － | － | － | － |  | － | － | － |
| SE | 14 | 7 | 7 | 7 | 2 | 1 | 2 | 8 | 8 | 78 | 44 | 56 | 8 | 6 | 5 | 5 | 20 | 67 | － | － | － | － | － | － | 6 | 4 | 8 | 12 | 17 | 59 | 100 |
| UK（E， <br> W\＆NI） | 129 | 143 | 160 | 169 | 176 | 439 | 2174 | 668 | 781 | 218 | 183 | 98 | 104 | 83 | 34 | 99 | 303 | 13 | 320 | 371 | － | 248 | 550 | － | － | － | － | － | － | － | － |
| UK （Scot．） | $\begin{aligned} & 703 \\ & 9 \end{aligned}$ | 7887 | 9712 | 11683 | 15658 | 22344 | 18783 | 13318 | 9710 | 9559 | 10024 | 8539 | $\begin{aligned} & 603 \\ & 3 \end{aligned}$ | $\begin{aligned} & 628 \\ & 4 \end{aligned}$ | $\begin{aligned} & 600 \\ & 3 \end{aligned}$ | 7722 | 8304 | 8658 | 7509 | $\begin{aligned} & 573 \\ & 0 \end{aligned}$ | － | $\begin{aligned} & 462 \\ & 2 \end{aligned}$ | $\begin{aligned} & 415 \\ & 4 \end{aligned}$ | － | － | － | － | － | － | － | － |
| UK （total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | $\begin{aligned} & 635 \\ & 3 \end{aligned}$ | $\begin{aligned} & 487 \\ & 0 \end{aligned}$ | $\begin{aligned} & 470 \\ & 4 \end{aligned}$ | $\begin{aligned} & 594 \\ & 3 \end{aligned}$ | $\begin{aligned} & 800 \\ & 5 \end{aligned}$ | 9296 | $\begin{aligned} & 1012 \\ & 7 \end{aligned}$ | 9735 | 7401 | 8092 |  |
| Total | $\begin{aligned} & 923 \\ & 5 \end{aligned}$ | 10209 | 12309 | 14505 | 17891 | 25176 | 23425 | 16859 | 13344 | 12338 | 12861 | 1104 8 | 852 3 | 898 7 | 842 4 | 1034 0 | 1063 2 | 1103 8 | 1006 7 | 819 0 | 776 0 | 645 9 | 639 3 | 763 3 | 969 0 | 1168 0 | 1362 0 | 1343 8 | $1115$ | 11739 | 11959 |

＊Preliminary．

Table 4．4．Continued．Nominal landings（ $\mathbf{t}$ ）of Anglerfish in the North Sea，as officially reported to ICES．
Central North Sea（4．b）
＊Preliminary

|  | ন্ন | No | $\stackrel{\oplus}{\circ}$ | す | 늑 | இั | 국 | ® | 욱 | 음 | 뭄 | N | 웅 | O | 승 | Ò O | 수N | 우N | Oio | $\circ$ N | 극 | N̈ | $\stackrel{m}{N}$ | $\stackrel{\rightharpoonup}{Z}$ | $\stackrel{\sim}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{\sim}{1} \end{aligned}$ | $\stackrel{\underset{\sim}{i}}{ }$ | $\stackrel{\infty}{\underset{\sim}{7}}$ | $\underset{\sim}{\underset{\sim}{2}}$ | ＊ | $\stackrel{*}{\text { N }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 357 | 538 | 558 | 713 | 579 | 287 | 336 | 371 | 270 | 449 | 579 | 435 | 180 | 260 | 207 | 138 | 179 | 181 | 134 | 124 | 111 | 131 | 135 | 213 | 196 | 251 | 168 | 155 | 249 | 239 | 284 |


|  | - | N్ | $\stackrel{ }{\underset{\sim}{2}}$ | す | 늑 |  | 축 | $\underset{\sim}{\circ}$ | 稌 | O- | Ö | No | Ò N্N | O | ㅅNㅇ | O- 무 | ì | 웅 | Oio | 0 <br>  | $\underset{\sim}{7}$ | $\underset{\sim}{\sim}$ | $\stackrel{m}{N}$ | $\underset{\sim}{\text { N }}$ | ํㅡㅊ | $\begin{aligned} & 0 \\ & \underset{\sim}{n} \end{aligned}$ | $\underset{\sim}{i}$ | $\stackrel{\infty}{\text { N }}$ | $\stackrel{\overbrace{}}{\mathrm{N}}$ | $\stackrel{*}{\sim}$ | * |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 345 | 421 | 346 | 350 | 295 | 225 | 334 | 432 | 368 | 260 | 251 | 255 | 191 | 274 | 237 | 276 | 173 | 237 | 248 | 194 | 286 | 301 | 192 | 334 | 369 | 584 | 565 | 411 | 533 | 339 | 322 |
| Faroes | - | - | 2 | - | - | - | - | - | - | - | - | 10 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| France | - | 1 | - | 2 | - | - | - | - | - | - | - | - | - | + | - | + | + | - | 3 | 6 | 2 | +- | +- | 1 | + | + | - | + | - | 2 | - |
| Germany | 4 | 2 | 13 | 15 | 10 | 9 | 18 | 19 | 9 | 14 | 9 | 17 | 11 | 11 | 9 | 14 | 12 | 22 | 17 | 21 | 17 | 10 | 10 | 17 | 23 | 18 | 14 | 26 | 27 | 16 | 22 |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | - | 1 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Netherlands | 285 | 356 | 467 | 510 | 335 | 159 | 237 | 223 | 141 | 141 | 123 | 62 | 42 | 25 | 31 | 33 | 61 | 58 | 36 | 46 | 53 | 61 | 41 | 72 | 88 | 120 | 166 | 111 | 310 | 226 | 354 |
| Norway | 17 | 4 | 3 | 11 | 15 | 29 | 6 | 13 | 17 | 9 | 15 | 10 | 12 | 22 | 16 | 12 | 24 | 15 | 21 | 10 | 11 | 11 | 26 | 8 | 9 | 16 | 41 | 36 | 22 | 42 | 17 |
| Spain | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | + | - | - | - | - | - | - | - | - | - |
| Sweden | - | - | - | 3 | 2 | 1 | 3 | 3 | 4 | 3 | 2 | 9 | 2 | 1 | 4 | 4 | 6 | 9 | - | - | - | - | - | - | 3 | 7 | 10 | 12 | 19 | 14 | 13 |
| UK(E, W\&NI) | 669 | 998 | 1285 | 1277 | 919 | 662 | 664 | 603 | 364 | 423 | 475 | 236 | 167 | 120 | 96 | 108 | - | 105 | 85 | 88 | - | 85 | 70 | - | - | - | - | - | - | - | - |
| UK (Scotland) | 845 | 733 | 469 | 564 | 472 | 475 | 574 | 424 | 344 | 318 | 378 | 210 | 241 | 138 | 88 | 98 | - | 142 | 108 | 125 | - | 115 | 72 | - | - | - | - | - | - | - | - |
| UK (total) | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 293 | - | - | - | 284 | 200 | 142 | 175 | 297 | 201 | 143 | 72 | 144 | 189 | 172 |
| Total | 2522 | 3053 | 3143 | 3445 | 2627 | 1847 | 2172 | 2088 | 1517 | 1617 | 1832 | 1244 | 847 | 851 | 688 | 683 | 749 | 769 | 651 | 615 | 764 | 714 | 546 | 820 | 985 | 1196 | 1107 | 823 | 1303 | 1066 | 1184 |

## Table 4．4．Continued．Nominal landings（ t ）of Anglerfish in the North Sea as officially reported to ICES．

Southern North Sea（4．c）
＊Preliminary．

|  | ন্ন | 국 | $\underset{ন}{\text { ® }}$ | G | 윽 | 윽 | 人 | が | 욱 | Oi | O-̇ | Nì | ÒN | ষ্ণ | NiO | O O | Nò | Oi ì | OiO | Oì | تָ | Nì | $\underset{\sim}{n}$ | $\underset{\sim}{\text { N }}$ | $\stackrel{n}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{N}{N} \end{aligned}$ | $\stackrel{N}{N}$ | $\stackrel{\infty}{\sim}$ | $\stackrel{\rightharpoonup}{\mathrm{N}}$ | ＊ | ＊ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 13 | 12 | 34 | 37 | 26 | 28 | 17 | 17 | 11 | 15 | 15 | 16 | 9 | 5 | 4 | 3 | 3 | 4 | 6 | 7 | 6 | 2 | 2 | 4 | 5 | 2 | 1 | 1 | 1 | 2 | 5 |
| Denmark | 2 | ＋ | － | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ | ＋ |  | ＋ | ＋ | － | － | ＋ | ＋ | ＋ | － | － | － | ＋ | － | ＋ | － | － | － |
| France | － | － | － | － | － | － | － | 10 | － | ＋ | － | ＋ | － | － | － | ＋ | ＋ | － | 1 | 1 | 1 | ＋ | ＋ | 1 | ＋ | 1 | ＋ | ＋ | － | － | ＋ |
| Germany | － | － | ＋ | ＋ | ＋ | － | － | － | － | － | － | － | － | － | － | ＋ | － | ＋ | ＋ | － | ＋ | － | ＋ | ＋ | ＋ | $+$ | ＋ | ＋ | － | ＋ | $+$ |
| Netherlands | 5 | 10 | 14 | 20 | 15 | 17 | 11 | 15 | 10 | 15 | 6 | 5 | 1 | － | 1 | － | 1 | 1 | － | 2 | 1 | 1 | 1 | 19 | 10 | 8 | 5 | 8 | 26 | 19 | 11 |
| Norway | － | － | － | － | ＋ | － | － | － | ＋ | － |  | － | ＋ | － | － | ＋ | － | － | 1 | － | － | － | － | 1 | ＋ | － | － | 1 | － | － | － |
| UK（E\＆W\＆NI） | 6 | 17 | 18 | 136 | 361 | 256 | 131 | 36 | 3 | 1 | － | － | 10 | 3 | － | ＋ | － | ＋ | 1 | 1 | － | － | 1 | － | － | － | － | － | － | － | － |
| UK（Scot－ land） | ＋ | ＋ | ＋ | 17 | ＋ | 3 | 1 | ＋ | ＋ | ＋ | － | － | － | 7 | － | ＋ | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| UK（Total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | －＋ | －＋ | ＋ | 1 | 1 | 1 | ＋ | 1 | 2 | 1 | 1 | ＋ | 1 | 2 | 1 | 2 |
| Total | 26 | 39 | 66 | 210 | 402 | 304 | 160 | 78 | 24 | 31 | 21 | 21 | 20 | 15 | 5 | 3 | 4 | 5 | 8 | 11 | 8 | 4 | 5 | 27 | 16 | 11 | 7 | 11 | 29 | 22 | 19 |

## Table 4．4．Continued．Nominal landings（t）of Anglerfish in the North Sea as officially reported to ICES．

Subarea 4 （North Sea）
＊Preliminary．／＾indicates landings assigned to Subarea 4 but not to a division．

|  | $\stackrel{7}{\text {－}}$ | ® | ¢ | 菏 | 告 | ঃঃ | － | $\stackrel{\circ}{\circ}$ | 条 | Oి | － | Oָ | O－N | ¢ | 適 | － | － | :্సి | Oio | 음 | $\underset{\sim}{\underset{I}{1}}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{\sim}$ | 茫 | $\stackrel{\sim}{\sim}$ | No | $\hat{\sim}$ | $\stackrel{\sim}{\sim}$ | ત̈뭄 | ＊ | $\stackrel{*}{\text { ® }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 372 | 559 | 595 | 753 | 607 | 323 | 357 | 389 | 286 | 476 | 594 | 459 | 190 | 265 | 211 | 141 | 181 | 185 | 140 | 131 | 116 | 133 | 137 | 217 | 200 | 253 | 169 | 156 | 249 | 243 | 290 |
| Denmark | 1599 | 1686 | 1293 | 1509 | 1027 | 1464 | 1489 | 1456 | 1496 | 1347 | 1540 | 1563 | 171 | 181 | 161 | 1587 | 1134 | 1308 | 1382 | 133 | 112 | 112 | 104 | 113 | 1331 | 2090 | 2567 | 2201 | 2202 | 1398 | 1462 |
| Faroes | 1 | － | 12 | 18 | 20 | － | 15 | 10 | 6 | － | 2 | 10 | 3 | 11 | 22 | 2 | － | － | 4 | － | － | － | － | － | － | － | － | － | － | － | － |
| France | 124 | 152 | 69 | 30 | 18 | 7 | 7 | 13 | 18 | 8 | 9 | 8 | 8 | 8 | 4 | 7 | 14 | 13 | 23 | 30 | 24 | 15 | 15 | 30 | 26 | 36 | 91 | 142 | 186 | 127 | 108 |
| Germany | 75 | 70 | 113 | 99 | 623 | 301 | 619 | 892 | 463 | 196 | 104 | 112 | 76 | 31 | 93 | 187 | 198 | 367 | 233 | 145 | 63 | 275 | 284 | 339 | 309 | 226 | 537 | 488 | 574 | 361 | 228 |
| Ireland | － | － | － | － | － | － | － | － | － | － | － | － | 1 | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Netherlands | 313 | 410 | 559 | 568 | 363 | 201 | 260 | 238 | 166 | 168 | 132 | 75 | 52 | 63 | 45 | 47 | 76 | 71 | 41 | 56 | 59 | 67 | 42 | 108 | 98 | 148 | 199 | 187 | 405 | 285 | 376 |
| Norway | 604 | 639 | 1227 | 1329 | 672 | 850 | 678 | 967 | 1236 | 1191 | 1227 | 938 | 781 | 102 | 896 | 1018 | 855 | 875 | 881 | 802 | 505 | 496 | 572 | 533 | 415 | 626 | 881 | 1267 | 1289 | 955 | 982 |
| Sweden | 14 | 7 | 7 | 10 | 4 | 2 | 5 | 11 | 12 | 81 | 46 | 65 | 10 | 7 | 9 | 10 | 26 | 76 | － | － | － | － | － | － | 10 | 11 | 18 | 25 | 36 | 72 | 112 |
| UK（E\＆W\＆N | 804 | 1158 | 1463 | 1582 | 1456 | 1357 | 2969 | 1307 | 1148 | 642 | 658 | 334 | 281 | 206 | 130 | 207 | 425 | 118 | 406 | 460 | － | 333 | 621 | － | － | － | － | － | － | － | － |
| UK（Scot－ | 7884 | 8620 | 1018 | 1226 | 1613 | 2282 | 1935 | 1374 | 1005 | 9877 | 1040 | 8749 | 627 | 642 | 609 | 7820 | 8476 | 8800 | 7617 | 585 | － | 473 | 422 | － | － | － | － | － | － | － | － |
| UK（Total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | 663 | 506 | 484 | 612 | 8303 | 9498 | 1027 | 9808 | 7545 | 8282 | 9604 |
| Total | $\begin{aligned} & 1179 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1330 \\ & 1 \end{aligned}$ | $\begin{aligned} & 1551 \\ & 9 \end{aligned}$ | $\begin{aligned} & 1816 \\ & 2 \end{aligned}$ | $\begin{aligned} & 2092 \\ & 0 \end{aligned}$ | $\begin{aligned} & 2732 \\ & 7 \end{aligned}$ | $\begin{aligned} & 2575 \\ & 7 \end{aligned}$ | $\begin{aligned} & 1902 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1488 \\ & 5 \end{aligned}$ | $\begin{aligned} & 1398 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1471 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1231 \\ & 3 \end{aligned}$ | $\begin{aligned} & 939 \\ & 0 \end{aligned}$ | $\begin{aligned} & 985 \\ & 3 \end{aligned}$ | $\begin{aligned} & 911 \\ & 7 \end{aligned}$ | $\begin{aligned} & 1102 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1138 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1181 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1072 \\ & 6 \end{aligned}$ | $\begin{aligned} & 881 \\ & 5 \end{aligned}$ | $\begin{aligned} & 853 \\ & 2 \end{aligned}$ | $\begin{aligned} & 717 \\ & 7 \end{aligned}$ | $\begin{aligned} & 694 \\ & 4 \end{aligned}$ | $\begin{aligned} & 848 \\ & 1 \end{aligned}$ | $\begin{aligned} & 1069 \\ & 1 \end{aligned}$ | $\begin{aligned} & 1288 \\ & 7 \end{aligned}$ | $\begin{aligned} & 1473 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1427 \\ & 4 \end{aligned}$ | $\begin{aligned} & 1248 \\ & 6 \end{aligned}$ | $\begin{aligned} & 1176 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1316 \\ & 2 \end{aligned}$ |
| Unallocated | $1224$ | － 1573 | － 2441 | － 2732 | － 5126 | $\begin{aligned} & 1108 \\ & 7 \end{aligned}$ | - 7540 | － 4999 | － 3166 | $2422$ | $2037$ | 600 | 676 | $\begin{aligned} & 133 \\ & 0 \end{aligned}$ | －579 | 1462 | 1561 | 1081 | 945 | 915 | 612 | 765 | 638 | 316 | 448 | 33 | 225 | －6 | －176 | 182 | －6 |
| WG esti－ | 1056 | 1172 | 1307 | 1543 | 1579 | 1624 | 1821 | 1402 | 1171 | 1156 | 1267 | 1171 | 871 | 852 | 969 | 9564 | 9823 | 1073 | 9781 | 790 | 792 | 641 | 630 | 816 | 1024 | 1285 | 1450 | 1428 | 1266 | 1158 | 1316 |
| mate | 6 | 8 | 8 | 0 | 4 | 0 | 7 | 7 | 9 | 4 | 7 | 3 | 4 | 3 | 6 |  |  | 2 |  | 0 | 0 | 2 | 6 | 5 | 3 | 4 | 8 | 0 | 2 | 1 | 8 |

Table 4．5．Nominal landings（ $\mathbf{t}$ ）of Anglerfish in Division 3．a，as officially reported to ICES．
＊Preliminary．

|  | － | 적 | $\stackrel{\text { ® }}{\underset{\sim}{2}}$ | す | 응 | இ冃 | ） | $\stackrel{\circ}{\circ}$ | 合 | ৪্ম | 밈 | ্ָ~ㅇ | O్N | ষ্ণ | 싱 | O్N | 윰 | :্స் | Oio | 음 | $\underset{\sim}{\underset{\sim}{1}}$ | N̈ | $\stackrel{\sim}{\sim}$ | 荡 | $\stackrel{\sim}{\sim}$ | $\stackrel{\square}{\sim}$ | $\stackrel{\text { Ni}}{~}$ | $\stackrel{\infty}{\sim}$ | ̈̈ | $\stackrel{*}{\sim}$ | $\stackrel{*}{\sim}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 15 | 48 | 34 | 21 | 35 | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － |
| Denmark | 493 | 658 | 565 | 459 | 312 | 367 | 550 | 415 | 362 | 377 | 375 | 369 | 215 | 311 | 274 | 227 | 255 | 287 | 344 | 270 | 251 | 307 | 298 | 309 | 336 | 389 | 526 | 597 | 692 | 600 | 678 |
| France | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | ＋ | ＋ | － | － | ＋ | 1 | － | － | － | － |
| Germany | － | － | 1 | ＋ | － | 1 | 1 | 1 | 2 | 1 | － | 1 | － | 1 | 1 | 2 | 1 | 1 | 1 | 1 | 2 | 1 | 1 | － | 1 | 2 | 1 － | 2 | 1 | 1 | 1 |
| Netherlands | － | － | － | － | － | － | － | － | － | － | － | － | 3 | 4 | 4 | 3 | 1 | 3 | － | 5 | － | － | － | 4 | 9 | 17 | 16 | 16 | 47 | 66 | 46 |
| Norway | 64 | 170 | 154 | 263 | 440 | 309 | 186 | 177 | 260 | 197 | 200 | 242 | 189 | 130 | 100 | 139 | 132 | 144 | 134 | 158 | 153 | 115 | 108 | 127 | 90 | 124 | 118 | 204 | 189 | 129 | 116 |
| Sweden | 23 | 62 | 89 | 68 | 36 | 25 | 39 | 33 | 36 | 27 | 46 | 55 | 71 | 73 | 79 | 54 | 44 | 51 | － | － | － | － | － | － | 42 | 53 | 81 | 95 | 100 | 71 | 71 |
| UK（Total） | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | － | ＋ | － | － | － | － | － |
| Total | 595 | 938 | 843 | 811 | 823 | 702 | 776 | 626 | 660 | 602 | 621 | 667 | 478 | 519 | 458 | 426 | 433 | 486 | 478 | 433 | 405 | 423 | 407 | 440 | 478 | 586 | 742 | 914 | 1029 | 866 | 912 |
| Unallocated | － | － | － | － | － | － | － | － | － | － | － | 288 | 252 | 197 | 174 | 189 | 168 | 187 | 79 | 109 | 116 | 63 | 65 | 78 | 66 | －5 | －911 | －22 | －34 | －22 | －30 |
| As used by WG | － | － | － | － | － | － | － | － | － | － | － | 379 | 226 | 322 | 284 | 237 | 265 | 299 | 399 | 324 | 289 | 360 | 342 | 362 | 412 | 591 | 751 | 936 | 1063 | 888 | 942 |

Table 4.6. Breakdown of WG estimates of commercial catches for 2020 and 2021 by main gear group and area in tonnes.

| 2020 | 3.a |  | 4 |  | $6 . \mathrm{a}$ |  | 6.b |  | Total |  | \% of Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards |
| Demersal trawl | 151 | 5 | 8296 | 86 | 4521 | 80 | 920 | 6 | 13888 | 178 | 73 | 56 |
| Nephrops trawl | 538 | 14 | 257 | 55 | 41 | 4 | 0 | 0 | 837 | 73 | 4 | 23 |
| Gillnets | 119 | 3 | 2434 | 40 | 32 | 1 | 817 | 5 | 3402 | 49 | 18 | 16 |
| Other/Not specified | 80 | 2 | 595 | 10 | 208 | 4 | 62 | 0 | 945 | 16 | 5 | 5 |
| Total | 888 | 22 | 11522 | 177 | 4763 | 88 | 1799 | 11 | 19072 | 316 | 100 | 100 |
| 2021 |  |  |  |  |  |  |  |  |  |  |  |  |
| Fleet | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards | Landings | Discards |
| Demersal trawl | 116 | 4 | 10423 | 95 | 4361 | 12 | 910 | 14 | 15810 | 125 | 78 | 50 |
| Nephrops trawl | 647 | 25 | 363 | 48 | 34 | 0 | 0 | 0 | 1044 | 73 | 5 | 29 |
| Gillnets | 107 | 5 | 1767 | 23 | 35 | 0 | 648 | 9 | 2557 | 37 | 13 | 15 |
| Other/Not specified | 72 | 5 | 616 | 8 | 9 | 0 | 35 | 0 | 732 | 13 | 4 | 5 |
| Total | 942 | 39 | 13169 | 174 | 4439 | 12 | 1593 | 23 | 20143 | 249 | 100 | 100 |

## Table 4.7. Total biomass estimates with confidence intervals and relative standard errors from the 2005-2021 SIAMISS-Q2 surveys.

| Year | Number of hauls | Number measured | Biomass (t) |  | Confidence Interval | RSE | Percentage Biomass in subarea 4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 |  |  | 38.617 | 23.479 | 53.755 | 20.0 | 48.27\% |
| 2006 |  |  | 40.985 | 34.478 | 47.492 | 8.1 | 53.49\% |
| 2007 | 156 | 1569 | 50.392 | 43.676 | 57.108 | 6.8 | 56.62\% |
| 2008 | 167 | 2219 | 53.546 | 42.421 | 64.671 | 10.6 | 55.51\% |
| 2009 | 206 | 1643 | 38.060 | 32.987 | 43.133 | 6.8 | 44.82\% |
| 2010 | 168 | 1280 | 42.279 | 30.429 | 54.129 | 14.3 | 51.90\% |
| 2011 | 153 | 1037 | 33.254 | 24.846 | 41.662 | 12.9 | 44.96\% |
| 2012 | 169 | 1461 | 36.325 | 29.704 | 42.946 | 9.3 | 41.59\% |
| 2013 | 93 | 984 | 38.395 | 31.020 | 45.770 | 9.8 | 37.04\% |
| 2014 | 106 | 1568 | 52.884 | 42.769 | 62.999 | 5.2 | 40.25\% |
| 2015 | 117 | 2198 | 67.915 | 58.782 | 77.047 | 6.9 | 43.66\% |
| 2016 | 108 | 2025 | 77.946 | 66.831 | 89.060 | 7.275 | 56.39\% |
| 2017 | 153 | 3265 | 87.896 | 74.222 | 101.569 | 7.937 | 53.47\% |
| 2018 | 142 | 2714 | 77.661 | 66.258 | 89.064 | 7.491 | 37.80\% |
| 2019 | 128 | 1860 | 58.575 | 46.189 | 70.962 | 10.789 | 40.49\% |
| 2021 | 137 | 1524 | 48.355 | 37.233 | 59.476 | 11.734 | 46.71\% |
| 2022 | 65 | 687 | 55.423 | 40.068 | 70.779 | 14.136 | 54.58\% |

## Table 4.8. Abundance and biomass estimates from the 2005-2021 SIAMISS-Q2 surveys by ICES subareas and divisions.

| Year | Month | Numbers (millions) |  |  |  |  | Biomass (kt) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | IVa | Vla | VIb | VI | Total | IVa | Vla | VIb | VI | Total |
| 2005 | November | 11.168 | 10.866 | 1.800 | 12.666 | 23.834 | 18.642 | 14.096 | 5.879 | 19.975 | 38.617 |
| 2006 | November | 12.844 | 10.459 | 3.174 | 13.633 | 26.477 | 21.921 | 12.175 | 6.889 | 19.064 | 40.985 |
| 2007 | November | 15.304 | 7.956 | 4.000 | 11.956 | 27.26 | 28.534 | 11.072 | 10.786 | 21.858 | 50.392 |
| 2008 | April | 12.613 | 7.718 | 3.952 | 11.67 | 24.283 | 29.721 | 14.383 | 9.442 | 23.825 | 53.546 |
| 2009 | April | 8.279 | 5.144 | 3.688 | 8.832 | 17.111 | 17.058 | 8.150 | 12.852 | 21.002 | 38.060 |
| 2010 | April | 7.366 | 5.161 | 3.131 | 8.292 | 15.658 | 21.944 | 11.59 | 8.745 | 20.335 | 42.279 |
| 2011 | April | 5.150 | 6.057 | 3.669 | 9.726 | 14.876 | 14.949 | 9.330 | 8.974 | 18.304 | 33.253 |
| 2012 | Abril | 5.432 | 4.961 | 5.135 | 10.096 | 15.528 | 15.106 | 9.213 | 12.005 | 21.218 | 36.325 |
| 2013 | October | 8.470 | 8.461 | 4.885 | 13.346 | 21.816 | 14.369 | 10.801 | 13.626 | 24.427 | 38.796 |
| 2014 | April | 17.553 | 16.096 | 6.488 | 22.584 | 40.136 | 21.284 | 16.633 | 14.967 | 31.60 | 52.884 |
| 2015 | April | 18.266 | 28.604 | 5.496 | 34.100 | 52.366 | 29.653 | 24.047 | 14.215 | 38.262 | 67.915 |
| 2016 | April | 21.648 | 14.383 | 4.538 | 18.922 | 40.569 | 43.956 | 18.273 | 15.717 | 33.99 | 77.946 |
| 2017 | April | 23.691 | 16.332 | 4.360 | 20.683 | 44.374 | 46.995 | 29.297 | 11.604 | 40.901 | 87.896 |
| 2018 | April | 11.819 | 13.528 | 6.240 | 19.768 | 31.586 | 29.353 | 22.350 | 25.958 | 48.308 | 77.661 |
| 2019 | April/May | 14.606 | 21.032 | 3.592 | 24.624 | 39.231 | 23.719 | 18.864 | 15.992 | 34.856 | 58.575 |
| 2021 | April | 17.371 | 8.608 | 3.048 | 11.656 | 29.027 | 22.587 | 12.74 | 13.027 | 25.767 | 48.355 |


| Year | Month | Numbers (millions) |  |  |  |  | Biomass (kt) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | IVa | Vla | VIb | VI | Total | IVa | Vla | VIb | VI | Total |
| 2022 | April | 13.259 | 10.283 | 3.048* | 13.331** | 26.591** | 30.252 | 12.142 | 13.027* | 25.170** | 55.423** |

*Value carried over from 2021 due to missing survey data in division 6b in 2022
**Total includes value carried over from 2021 due to missing survey data in division 6b in 2022

Table 4.9. Northern Shelf anglerfish mean standardised harvest rates of catch numbers and biomass 2008-2021.

| Year | Mean standardised harvest rate - Number | Mean standardised harvest rate - Biomass |
| :---: | :---: | :---: |
| 2007 | 1.019557 | 0.962803 |
| 2008 | 1.396944 | 1.010032 |
| 2009 | 1.117361 | 1.304825 |
| 2010 | 1.307811 | 0.930318 |
| 2011 | 1.08548 | 1.259594 |
| 2012 | 1.28022 | 1.041019 |
| 2013 | 1.058155 | 1.001878 |
| 2014 | 0.727179 | 0.938519 |
| 2015 | 0.62145 | 0.761601 |


| Year | Mean standardised harvest rate - Number | Mean standardised harvest rate - Biomass |
| :--- | :--- | :--- |
| 2016 | 0.861419 | 0.842697 |
| 2017 | 0.760659 | 0.824041 |
| 2018 | 0.908121 | 0.908698 |
| 2019 | 0.855643 | 1.213976 |
| 2021 | 0.855643 | 1.213976 |



Figure 4.1. Northern Shelf anglerfish. Officially reported landings by ICES area (1973-2021).


Figure 4.2. Trends in nominal international fishing effort (kW*days at sea) in North Sea and II (EU) (left) and West of Scotland (right) collated by STECF for the Evaluation of Fishing Effort Regimes in European Waters (STECF, 2017).


Figure 4.3a. Percentage of total landings weight by fleet and country in 2021; Subarea 4.


Figure 4.3b. Percentage of total landings weight by fleet and country in 2021; Division 6.a.


Figure 4.3c. Percentage of landings weight by fleet and country in 2021; Division 6.b.


Figure 4.4. WGCSE Landed numbers ('00 thousands) at-length (cm) 2002-2021.

## Stock size indicator



Figure 4.5. SIAMISS-Q2 estimates of total biomass with $95 \%$ confidence intervals for subareas 4 and 6 combined, 2005-2021. The horizontal orange lines indicate the average of the most recent two years, and the previous two years (with 2020 not available and hence excluded from the average). The dashed line represents the index trigger value.


Figure 4.6. SIAMISS-Q2 estimates of total abundance (left) and biomass (right) of anglerfish for the Northern Shelf 20052022 provided for ICES Subarea 4a, Division 6.a and Division 6.b.


Figure 4.7. SIAMISS-Q2 estimates of total numbers (millions) and weight ( t ) at-length ( cm ) for subareas 4.a and 6.a, 2022.


Figure 4.8. SIAMISS-Q2 estimates of total numbers (millions) at-length (cm) (left) and estimates of total biomass (kt) at-length (cm) (right) for subareas 4.a (blue)-c and 6.a (yellow)-b (red) combined, 2007-2022.


Figure 4.9. Percentage of SIAMISS-Q2 total biomass, with confidence intervals, estimated to be in subareas 4.a-c compared with subareas $4 . a-c$ and $6 . a-b$ combined. The full grey line represents the average of these percentages over the time-series (2005-2022) (48\%). The dotted grey lines represent the percentage of TAC allocated for subareas 4.a-c compared to the total of the TAC for subareas 4.a-c and 6.a-b, ( $67 \%$ in 2005-2010, 64\% in 2011-2022).


Figure 4.10. Northern Shelf anglerfish harvest rate 2008-2021 (mean standardised WG catch total numbers of biomass / SIAMISS-Q2 total numbers or biomass).

## Length indicator



Figure 4.11. Anglerfish in subareas 4 and 6 and in Division 3.a. Length indicator (mean length of fish in the catch above the estimated length of first capture (Lc) normalised to the MSY proxy reference length $\mathbf{L F}=\mathbf{M}$.). The exploitation status is below FMSY proxy when the indicator ratio value is higher than 1 (shown by a blue line).


Figure 4.12. Survey indices of mean weight (g) per hour from SWC-IBTS-Q1 (blue) in 6.a, SWC-IBTS-Q4 (red) in $6 . a$ and Rockall (red) in 6.b.


Figure 4.12. Survey indices of mean weight (g) per hour from NS-IBTS-Q1 (brown) and NS-IBTS-Q3 (orange).


Figure 4.13. ICES landings of anglerfish 1991-2021 (top) and ICES discards of anglerfish 2007-2021 (bottom) in subareas 4 and 6 and in Division 3.a.


Figure 4.14. Numbers of anglerfish per km² observed by SIAMISS surveys 2016-2022.


Figure 4.15. Weight of anglerfish (kg) per km² observed by SIAMISS surveys 2016-2022.

## 3 Cod in Division 6.a

### 3.1 Introduction

The last benchmark for this stock was carried out in February 2020 (ICES, 2020). This resulted in a change of assessment method (TSA to SAM), inclusion of revised catch data from 2003 onwards and updated biological parameters.

The assessment presented here contains a number of deviations associated with the catch estimation process and input data to the approaches agreed at the benchmark and documented in the Stock Annex:

Processed UK VMS data have not been submitted in accordance with the ICES VMS data call deadline and hence were not available ahead of WGCSE. Furthermore, data access issues between UK administrations mean that raw UK VMS data are not directly available for the estimation of Scottish area misreported landings. This applies to data from 2021. Therefore, instead of using these data to estimate area misreported landings (as agreed at WKDEM, ICES, 2020a), the WG has again had to make use of estimates provided by Marine Scotland Compliance (MS-C, which were used by the WG prior to the 2020 benchmark). VMS data for 2019 and 2020 became available just prior to the WG meeting and hence could not be used to revise the previously used MS-C estimates. The MS-C estimates are used for 2019-2021.

In 2021, Scottish observer sampling from the Nephrops trawl fleet (OTB_CRU) was extremely limited due to COVID-19 disruption. This has resulted in an underestimate of total discards and unreliable estimates of catch numbers and mean weights-at-ages 1 and 2 which as a consequence have been excluded from the assessment. Sensitivity analysis suggests this has had minimal impact on the assessment (See Section $3.3 \& 5.4$ for further details and sensitivity analysis).

Due to vessel breakdown in 2022, the Scottish Q1 West Coast survey could not be carried out and hence there was no intermediate year survey included in this year's assessment. Sensitivity analysis suggests this has had minimal impact on the assessment (See Section $3.3 \& 5.4$ for further details and sensitivity analysis).

### 3.2 General

### 3.2.1 Advice

This stock has had zero catch advice since 2004. In recent years, this advice has typically been issued on a biennial basis.

### 3.2.2 Stock definition and the management unit

The general conclusion from recent workshops on cod stock ID in Division 6.a and the neighbouring North Sea (WK6aCodID; ICES, 2022 \& WKNSCODID; ICES, 2020b) was that the current assessment units are not consistent with the stock structure.

WK6aCodID concluded that the available evidence supported a hypothesis of multiple overlapping subpopulations in Division 6a related to the Dogger genetic lineage (with linkage between 4 a and 6 a ) with a separate subpopulation in the Clyde, associated with the Celtic genetic lineage and for which evidence for a link to Division 7a was presented.

The non-Clyde part of Division 6a is considered likely to consist of separate inshore and offshore Dogger subpopulations, but given the very limited data from parts of Division 6a, there remains uncertainty regarding the spatial extent of these. Genetic evidence for a link with the North Sea Dogger unit was supported by tagging data showing mixing between the northern part of a and the northwestern part of 4a in the North Sea.

The Clyde population, genetically associated with the Celtic unit, is different to elsewhere in Division 6a in terms of otolith microchemistry and demographics (maturity and SSB trends). This is confirmed by tagging data showing Clyde cod to be largely resident within the Clyde, with no mixing with other 6a subpopulation and limited exchange with the Irish Sea.

WK6aCodID considered it highly unlikely that it would be possible to collate sufficient appropriately disaggregated data to enable a separate Clyde stock assessment to be conducted in the near future. However, given the currently very minor contribution of Clyde cod catches to the overall catches from Division 6a, the impact of retaining the Clyde cod within a meta-population stock assessment (as part of the Inshore Dogger subgroup within 6a) is likely to be minor and was therefore recommended as a practical way forward for the short to medium term.

Within the North Sea, WKNSCODID (ICES, 2020b) concluded that there were separate Viking (northeast North Sea) and Dogger (remaining North Sea) genetic populations (with boundaries agreed), with the northern offshore component of Division 6a considered likely to be part of the latter.

A process for developing a combined spatial assessment for North Sea and West of Scotland cod, accounting for the substock structure is underway, and will conclude in 2023.

The management unit is ICES Divisions 6.a plus EU and international waters of Division 5.b to the east of $12^{\circ} 00^{\prime} \mathrm{W}$. Prior to 2009, the TAC was set for ICES subareas 6, 12 and 14 plus Subdivision 5.b.1.

## Recent management

The minimum conservation reference size of cod for human consumption in this area is 35 cm .
From 2012 to 2018, the TAC for cod in Division 6.a was set to zero with allowance for a bycatch of cod to be landed provided that it does not comprise more than $1.5 \%$ of the live weight of the total catch retained on board per fishing trip. From 2015, this provision was not allowed for catches subject to the landing obligation.

With the full implementation of the landing obligation in 2019 for fisheries catching cod, a bycatch TAC of 1735 t was set to allow mixed fisheries with a bycatch of cod to continue. In 2020, this TAC has been reduced to 1279 t . The agreed TAC has remained at this level in 2021 and 2022, although the quota share has changed (as agreed under the EU-UK Trade and Cooperation Agreement). A breakdown of these TACs by country can be found below.

## TAC 2020

| Species: | Cod Gadus morhua |  | Zone: | 6a; Union and international waters of 5 b east of $12^{\circ} 00^{\prime} \mathrm{W}$ <br> (COD/5BE6A) |
| :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 2 (1) | Analytical TAC |  |
| Germany |  | 19 (1) | Article 8 of this Regulation applies |  |
| France |  | 203 ( ${ }^{\text {1 }}$ ) | Article 3 of Regulation (EC) No 847/96 shall not apply |  |
| Ireland |  | $284{ }^{(1)}$ | Article 4 of Regulation (EC) No 847/96 shall not apply |  |
| United Kingdom |  | $771{ }^{(1)}$ |  |  |
| Union |  | 1279 (1) |  |  |
| TAC |  | 1279 (1) |  |  |

${ }^{( }{ }^{1}$ ) Exclusively for by-catches of cod in fisheries for other species. No directed fisheries for cod are permitted under this quota.

## TAC 2021



## TAC 2022



Technical measures applicable to the West of Scotland, including those associated with the cod recovery plan in force up to 2008 (Council Regulation No. 423/2004), the cod long-term management plan in force from 2009 (Council Regulation No. 1342/2008) were amended by Council Regulation No. 1243/2012. The management plan was further amended in 2016 by Council

Regulation (EU) 2016/2094 to cover the transitional period in which preparations are ongoing towards multiannual plans for multispecies fisheries. In 2018, the cod management plan was discontinued. Cod in Division 6.a is not included as a named target species in the EU multiannual plan for Western Waters (Council Regulation (EU) 2019/472).

### 3.2.3 The fishery in 2021

The table of official landings statistics is given in Table 3.1 and Figure 3.1. Official landings increased in 2021 ( 1209 t ) compared to $2020(983 \mathrm{t})$. Note that updates to official landings data for 2019 associated with national GDPR clauses means that data for this year are now incomplete. In 2021, over $75 \%$ of the official landings were reported by UK vessels, approximately $15 \%$ by France with smaller amounts declared by Ireland and Spain. The majority of reported cod landings in Division 6.a are now taken in the far north of the area. In 2021, officially reported BMS (below minimum size) landings of cod in Division 6.a were less than half a tonne.

Due to restrictive TACs, seasonal/spatial closures of the fishery, and effort restrictions based on bycatch composition, the likelihood of misreporting and underreporting of cod in the past is considered to have been high. Underreporting is considered to have been reduced to low levels following the introduction of legislation in Ireland and the UK in 2006. However, area misreporting of cod landings from Division 6.a into Division 4.a (i.e. caught in Division 6.a., but declared in Division 4.a) and to a lesser extent Division 5.b, by the Scottish fleet is believed to occur. The UK legislation introduced in 2006 is also believed to be responsible for a significant increase in discards starting in 2006. Following the full implementation of the landing obligation (2019 onwards) for fisheries catching cod and the availability of a bycatch TAC rather than a $1.5 \%$ bycatch allowance, discard rates have been much reduced since 2019 although area misreporting continues to occur (albeit at an apparently lower level in recent years).

### 3.3 Data

## Catch data

Area-misreported landings by the Scottish fleet have been considered to represent a considerable proportion of the total landings. One of the main focuses of the 2020 benchmark was deriving an objective approach for estimating area misreported landings based on an analysis of VMS data linked to daily logbook landings (WD 4.4, ICES, 2020a) rather than using estimates provided by Marine Scotland Compliance (MS-C) based on fishery observations and expert judgement (as used by the WG prior to the 2020 benchmark). However, UK VMS data for 2019 onwards have not been submitted in accordance with the VMS data call deadline and hence have not been available prior to WGCSE. Therefore, as in 2020 and 2021, the WG again had to revert to making use of area misreported landings estimates provided by MS-C. Figure 3.2 and Table 3.2 shows the time-series of estimates of area misreported landings (which come from the UKS large mesh demersal trawl fleet) alongside reported landings for Division 6.a. Total estimated area misreported Division 6.a cod landings in 2021 are 49 t . This represents a decline both in total weight of area-misreported landings and also a reduction in the proportion relative to total landings remains similar ( $<5 \%$ ). These landings are largely reported into Division 4.a, but assumed to actually be taken in Division 6.a. It is not clear why this sharp reduction has taken place. The approaches to identify area-misreporting used by MS-C have not changed since 2020. One explanation could be that an increase in quota share for 6 (for the UK) coupled with a highly restrictive N Sea cod quota have meant that there is less need/opportunity to area misreport across the 4 degrees west line.

The landings uploaded into InterCatch are shown in Figure 3.3 by métier and country, and discard proportions by weight shown in Figure 3.4. The French OTB_DEF $\geq 120$ métier is the largest metier with unsampled landings and represents $9 \%$ of the total landings in 2021.

In 2021, fishery sampling continued to be disrupted by the COVID-19 pandemic. Sampling of both landings and discards from the main fleet (Scottish OTB_DEF) remained at around half the number of trips in pre-pandemic years. While the number of samples was lower for this fleet, samples were available from both sources (landings and discards) with reasonable seasonal coverage. The most significant impact of the reduced sampling was on the number of samples and seasonal coverage of discard samples from the Nephrops trawl fleet (Scottish OTB_CRU). The number of samples from this fleet was around $25 \%$ of typical levels (four trips). This fleet usually has a discard rate of almost $100 \%$, but none of the sampled trips caught (and discarded) any cod. Given that there have been no changes in selectivity devices used in the fishery that would result in a reduction in discards, it was considered that these trips were unlikely to be representative of the fishery as a whole and the estimates of zero discards were not uploaded into InterCatch.
Following an analysis of Scottish catch sampling data conducted at WKDEM (ICES, 2020a), it was agreed that for the purposes of allocated age compositions and discard rates, the area-misreported landings should be considered as 'sampled' landings and treated as part of the Scottish demersal trawl fleet. This is in contrast to previous assessment WGs where the area-misreported component was considered un-sampled and were assumed to have zero discards and landings age compositions consistent with the total sampled landings (i.e. all countries).

Due to the lack of discard sampling from the OTB_CRU fleet in 2021, discard proportions and landings and discard age distributions for all unsampled fleets were assigned from the only sampled fleets (Scottish \& Irish OTB_DEF fleets) within InterCatch (representing a deviation from the Stock Annex). Allocated discard rates using this approach are shown in Figure 3.5 and estimated total catch by metier in Figure 3.6. The final mix of numbers-at-age from sampled and unsampled landings and sampled and raised (un-sampled) discards is given in Figure 3.7. An extremely small amount ( $<0.5 \mathrm{t}$ ) of below minimum size (BMS) landings was also reported, but is not shown. There is a noticeable lack of age 1 fish in the catch in 2021. Figure 3.8 shows the breakdown of catch numbers-at-age by fleet (OTB_DEF/OTB_CRU) and catch category (landings/discard) in 2021 compared to 2018-2020 (note that 2018 is pre-LO). The OTB_CRU fleet typically catches younger individuals than the OTB_DEF fleet and a significant proportion of the catch numbers-at-age 1 (and to a lesser degree age 2) in 2018-2020 are taken by the OTB_CRU fleet (grey colour in Figure 3.8). These are absent from the data in 2021. As a result, the WG concluded that total discards and catch numbers-at-age 1 and 2 for 2021 were likely to be underestimated due to the lack of samples from the OTB_CRU fleet.
Sampling levels (number of trips) by country are given below and compared to 2019. A limited number of Northern Irish samples are also available in 2020 and 2021. Sampling of the Scottish OTB_DEF landings has been quite poor in the recent past. The small sample sizes (which include a few very large fish with high raising factors) can result in a very high sum of products (SOP, landings-at-age x weight-at-age) for this fleet in some years.

|  | Scotland |  |  | Ireland |  |
| :--- | :--- | :--- | :--- | :--- | :---: |
| 2019 | Year | Demersal trawl (OTB_DEF) | Nephrops trawl (OTB_CRU) | Total | Total |
|  | Observer | 19 | 1 | 20 | 21 |
| 2020 | Landings | 22 | 18 | 40 | 28 |
| 2021 | Observer | 10 | 4 | 10 | 24 |
|  | Landings | 11 | 0 | 14 | 5 |
|  | Observer | 9 | $0^{\wedge}$ | 28 |  |

$\wedge$ Four trips sampled with zero discards. Not used due to low confidence in estimates due to low sample size.

The WG estimates of total landings and discards are given in Table 3.2 and shown in Figure 3.9. The total discard proportion by weight is shown in Figure 3.10, and while this has increased somewhat in 2021 ( $\sim 35 \%$ ) compared to 2019 (9\%), it remains well below the previous 3-year average. (2016-2018) when the discard proportion was estimated to be in excess of $70 \%$ of the total catch.

In contrast to the period 2006 to 2018 when there was substantial highgrading and discarding occurring (to some degree) over all age classes, during the years 2019 to 2021, discarding is mostly limited to ages 1 and 2 (and to a lesser degree age 3) (Figures 3.11 and 3.12). In 2021, there is a reduction in discard proportion at age 1 compared the previous ten years. However, given the lack of Nephrops discard sampling data (and the underestimate of discards at age 1), it seems unlikely that this apparent reduction is a true indication of changes in the fishery.

## Age-compositions and weights-at-age

Raised landings numbers-at-age and discard numbers-at-age are given in Tables 3.3 and 3.5 respectively and total catch numbers-at-age in Table 3.7.

Annual mean weights-at-age in landings, discards and catch are given in Tables 3.4, 3.6 and 3.8. Figure 3.13 shows the mean weights-at-age in the landings and discards. The mean weight of age two and three fish in the landings increased since the mid-2000s in line with the increase in highgrading which occurred at these ages. Other age classes show fluctuations with a long-term downward trend particularly for ages 5 and above. Values at older age are noisy, particularly in recent years (most likely due to low sampling levels). Mean weight-at-age in the discards shows no real trend between 2006 and 2018. In 2020, there is a decline in mean weight-at-age in both the landings and discards at age 1 . While the 2020 estimates remain well within historical values (and are not extreme), the lower discard mean weight could potentially be due to the lack of discard samples from quarter 2 onwards (i.e. after individuals have grown) from the Nephrops fleet which is typically the most important fleet for age 1 discards. (See above for COVID-19 samples disruption). The reason for the decline in mean weight-at-age 1 in the landings is harder to explain and potentially is noise related to low sample sizes or increased retention of smaller fish (due to the LO). In 2021, there is a substantial increase in mean discard weight (and subsequently catch weight) at age 1 and 2. Closer inspection of mean discard weights-at-age by fleet (Scottish data), Figure 3.14, suggests that this may be due to a lack of samples from the OTB_CRU fleet as this fleet generally catches smaller individuals (at age) than the OTB_DEF fleet. In addition, a number of very large age 1 fish were recorded in 2021 OTB_DEF samples which contribute
to the high value. The WG agreed that these discard (and catch) mean weights-at-age in 2021 are therefore likely to be biased estimates.

## Biological data

Given the trends in observed mean weights, WKDEM proposed the use of a temporally varying natural mortality would be more appropriate. The catch weights show high interannual variability (Figure 3.13) and therefore it was agreed to use smoothed catch weights as stock weights and then use these with the Lorenzen (1996) function with the 'natural' parameters to obtain natural mortality (WD 4.3, ICES, 2020a).

To derive the stock weights, a GAM is fitted to mean catch weights-at-age (Figure 3.14). Refitting the GAM each year results in typically minor revisions to stock weights used to estimate SSB between assessment years (and also natural mortality, WD 4.3, ICES, 2020a). Including the biased estimates of mean catch weight-at-age 1 and 2 in 2021 in the GAM has a significant impact on the estimated stock weights-at-age. The WG therefore agreed that these values should be excluded from the smoothing process for estimating stock weights, and stock weights-at-age 1 and 2 for 2021 should be set equal to the estimated values for 2020.

The catch mean weight-at-age 2 in 2019 remains a substantial outlier. At WGCSE 2020, the sampling data for 2019 were scrutinized in detail and the estimate could not be attributed to a particularly anomalous or influential sample and therefore the datapoint was considered valid (See ICES, 2020 for further details).

At all ages there is a general downward trend in catch weights (and hence stock weights) over time although with an apparent recent increase at ages 3 and 4 . This results in increases in natural mortality, although at most ages the scale of this increase is very small (Figure 3.15). Stock weights and natural mortality are given in Tables 3.9 and 3.10.

The maturity ogive was also updated at WKDEM. An analysis of Scottish survey data (following the approach advocated by ICES, 2008) indicated a proportion of individuals at age 1 to be mature, but no temporal trend in maturity. A new ogive was therefore used for the full time-series (WD 4.2, ICES, 2020a).

| Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| WGCSE 2019 WKDEM/WGCSE 2020 onwards | 0 | 0.52 | 0.86 | 1.0 | 1.0 | 1.0 | 1.0 |

## Survey data

All available survey data are given in Table 3.11, with the data used in the assessment highlighted in bold. Survey descriptions are given in the stock annex. Since the inter-benchmark in 2019 (IBPCod6.a), the assessment makes use of three quarter four surveys (one of which is no longer current) and two quarter one surveys (one of which is discontinued). Survey indices for the two current Scottish surveys (UK-SCOWCGFS- Q1 and UK-SCOWCGFS- Q4) are provided with an estimate of variance.

The CPUE by survey haul for recent years for the two Scottish surveys (UK-SCOWCGFS- Q1 and UK-SCOWCGFS- Q4) are shown in Figure 3.16. Both surveys show mostly zero returns over latitudes between 56 degrees N and 58.5 degrees N . This pattern has been consistent in surveys since 2007. The Scottish surveys have highest catch rates to the north of 59 degrees N , in and around the closed area although these seem to have reduced in recent years (coincidental with a
reduction in the size of the area closed to fishing, Figure 3.16). South of 56 degrees N, the Q1 surveys catch mostly young cod in the Clyde region. Occasional very large hauls associated with apparent aggregations of older cod (typically age 3 and above) have a significant impact on the survey indices and their variance estimates. In 2017, the indices for age four, five and six cod in the quarter one survey show particularly high uncertainty due to a single very large haul (Figure 3.16) of large cod with most other stations having very low or zero values. In 2018 (in the same survey), there were no large hauls and therefore the estimated variance is low. In 2019, the quarter one survey shows very low catch rates of ages $>1$ across the survey area, but relatively high catch rates (compared to recent years) of age 1 fish.

The quarter four survey estimates also have substantial uncertainty. This is particularly apparent in the 2016 survey with a CV of over $70 \%$ at age 4, and to a lesser degree in the 2018 survey with two hauls catching large numbers of individuals aged 4 to 6 and very low catches elsewhere, resulting in CVs of around $60 \%$ for these ages in this year.

Due to vessel breakdown the Scottish Q1 survey was not carried out in 2022.
A series of inshore and offshore Scottish industry-science surveys, known as the West Coast Demersal Fish (WCDF) project were conducted between December 2013 and November 2014. The initiative, funded by the Scottish Government and the European Fisheries Fund, was a joint venture between Marine Scotland Science and the Scottish Fishermen's Federation with the aim of improving the understanding of the current state of demersal stocks to the West of Scotland. The surveys show a broadly similar distribution to the UK-SCOWCGFS- Q1 and UK-SCOWCGFS-Q4 with bigger fish and increased abundance inside the Windsock compared to outside. Biomass estimates from these surveys and from the SIAMISS (anglerfish survey) were presented to WKDEM, but were considered too uncertain to provide useful information for the stock assessment.

### 3.4 Stock assessment

This assessment uses a SAM run as outlined in the stock annex. Exploratory analysis of the input catch and survey data are also carried out.

## Data screening

Log catch (landings + discards) numbers-at-age over time (Figure 3.17) show good tracking of strong and weak cohorts historically. These signals become less apparent and more noisy after 2010, potentially due to low sampling levels and/or ageing errors. There is however, a clear indication of increasing numbers of older fish appearing in the catch since this time, which would be consistent with a reduction in fishing mortality. Catch curves from commercial catch-at-age data are also shown in Figure 3.17. Although the data are noisy, there is some evidence of a flattening off of the catch curves in recent years compared to those of the cohorts spawned in the late 1990s. Figure 3.17 shows that the $\log$ catch numbers-at-age 1 are by far the lowest of the time-series which supports the view that these are potentially biased (due to lack of OTB_CRU discard sampling).

A plot of log catch curve gradients derived from commercial catch data over different age ranges is shown in Figure 3.18. Here too there is some evidence of a decreasing mortality in recent years. (Note that these exploratory catch data plots are based on reported landings and discards and will be influenced in part by underreporting of landings in the 1990s and early 2000s).

Figure 3.19 shows the mean standardised catch-at-age by proportion (number). It shows good tracking of the strong cohorts as recently as the 2005 year class which shows well up to age 4.

More recently the data become rather noisy and since 2018, the proportion of the catch-at-age at age four and above are very high. These observations are not supported consistently by above average values at younger ages of the same cohort. Potentially this could be associated with a slight change in the distribution of the fishery and access to a previously closed area (illustrated in Figure 3.16) where a significant proportion of the older fish are located (Figure 3.16), however recent VMS data are unavailable and hence this hypothesis cannot be substantiated.

Figure 3.20 shows the log mean standardised indices from the ScoGFS-WIBTS-Q1 survey by year and by cohort. The early part of the time-series appears to track the cohorts relatively well with no obvious year effects. However, in later years the indices become noisier and there is some evidence of year effects in the survey. The survey ended in 2010. Figure 3.21 shows log catch curves for the ScoGFS-WIBTS-Q1 survey. It shows a strong "hook" at the younger ages (lower catchability), with abundance-at-age two often higher than at-age one. In later years survey abundance also shows increases from age 2 to age 3 in the same year class and the survey's ability to track recent cohorts seems poor relative to the 1990s and early 2000s. The survey scatterplots (Figure 3.22) show some consistency in the estimates of year-class strength across age classes (particularly the younger, adjacent ages), although less so at older ages. There is no trend in the log catch curve gradients derived from this survey that would be consistent with a change in mortality (Figure 3.23) for any of the age ranges considered.

Figure 3.24 shows the log mean standardised indices by cohort and year from the ScoGFS-WIBTS-Q4 survey. The survey shows reasonable tracking of cohorts at ages one to three and no particular evidence of year effects. This is also evident in the survey scatterplots which show reasonable correlation at younger ages (Figure 3.25). This survey catches very few fish at ages five and above.

Figure 3.26 shows the log mean standardised indices by cohort and year from the IRGFS-WIBTSQ4. The $\log$ mean standardised indices plot shows consistent signals at ages 1 and 2 early in the time-series with no obvious year effects. The scatterplots (Figure 3.28) also show reasonable consistency between ages one and two, but the tracking at older ages is less strong. The data cover too few age classes sufficiently well to give an indication of trend in mortality through catch curve gradients (Figure 3.27).

Figure 3.29 shows log mean standardised indices by cohort and year from the UK-SCOWCGFSQ1. Cohorts tracking within this survey is inconsistent and there is some evidence of survey year effects (2015, 2017 and 2019, particularly for older ages). There appeared to be a general increase in the catch rates of older ages over time to 2017 (four and above), but no equivalent increase in the catch rates of younger ages (from the same cohort). These declined significantly in 2018 and 2019.

The log catch curves from the UK-SCOWCGFS- Q1 are also very noisy (Figure 3.30) and typically do not show a decline as the cohort ages. The survey scatterplots show that even the catch rates of successive age classes (within the same cohort) show weak positive correlation (Figure 3.31).

Figure 3.32 shows log mean standardised indices by cohort and year from the UK-SCOWCGFSQ4. There is some evidence of cohort tracking, but this is not consistent over time or ages and this is also apparent in the survey scatterplots shown in Figure 3.34. Figure 3.33 shows the log catch curves from the UK-SCOWCGFS-Q4 which are noisy and difficult to interpret given the short time-series and missing year of survey data.

Overall, information on mortality trends from all survey-series (including the ScoGFS-WIBTSQ1) appears to be fairly poor due to the generally high variability and large CVs (ranging from $30 \%$ to $75 \%$ depending on age-class) for the two current Scottish surveys.

Figure 3.35 shows a comparison (between surveys) of log mean standardised survey indices at age over time (mean standardised over the common year range of all three surveys). The two

Scottish surveys show reasonable consistency over ages two to four, despite being noisy. The Irish survey also shows reasonable agreement at age two. At older ages (in the Scottish surveys), the general trends are similar, but show different interannual variations.

The inter-benchmark in 2019 agreed that all five surveys should be included in the final assessment (and this was followed at WKDEM in 2020), the basis being that the additional surveys show reasonable internal consistency and in addition, some between survey consistency. It was considered that the Irish survey could provide an additional indicator of year-class strength and could be useful as it covers the period during which there is a break in the Scottish survey indices. The lack of spatial coverage of this survey (only the southern part of Division 6.a) was deemed less important given the index is only being used to provide information on the younger ages.

## Final assessment

The SAM configuration file for the final assessment model run is given in Table 3.12. To summarise the main features:

- Fishing mortality at ages 4 and above are assumed equal (See \# Coupling of the fishing mortality states, Table 3.12).
- Survey catchabilities are mostly freely estimated for each age with the exception of the two oldest ages (i.e. no survey catchability plateau assumed). The exception to this is the WIBTS.Q1 for which all catchabilities are independently estimated.
- Catch observation variance parameters are allowed to differ for age 1 and age 7+ while other age groups are coupled (\# Coupling of the variance parameters for the observations). To allow for greater uncertainty in the catch data for 2006 onwards (when the fishery changes from being a landings fishery to largely discards), the estimated catch observation error standard deviation is doubled for 2006 onwards (based on inspection of the one step ahead residuals).
- Survey observation variance parameters differ between surveys but are coupled for all age groups within a survey.
- Recruitment is modelled as a random walk.
- A catch scaling factor is estimated for 1995-2006 when underreporting of landings was considered significant.
- Fishing mortality across ages is modelled with $\operatorname{AR}(1)$ and process variance parameters coupled across all ages with the exception of age 1. Process variance in stock numbers-at-age were assumed coupled with the exception of age 1 (the age at recruitment).

Input data are derived as agreed at the 2020 benchmark with a number of exceptions:

- the use of MS-C estimates of area-misreported landings for 2019-2021 rather than estimates from VMS data (as per assessment WGs in 2020 and 2021). A comparison of VMS estimates and MS-C estimates carried out at WKDEM suggested VMS estimates were generally lower with some correlation between the two sets of estimates.
- The lack of an intermediate year survey (Q1 2022 data missing) due to vessel breakdown.
- The exclusion of age 1 and age 2 catch numbers-at-age due to concerns over bias (likely underestimated) due to lack of discard sampling from the OTB_CRU fleet.

Sensitivity analyses have been conducted to explore the potential impacts of the missing survey data and the exclusion of the catch data. (No further sensitivity analysis was carried out regarding the use of MS-C data since this approach has now been utilised at the past three assessment WGs).

Figure 3.36 shows a retrospective sensitivity analysis to the exclusion of the intermediate year data i.e. comparing previous years' assessments with and without these data included. While
the intermediate year estimate of recruitment is sensitive to the inclusion (or not) of the intermediate year survey, the exclusion of the data does not result in major historical revisions to either recruitment, SSB or F. In addition, the analysis also suggests that the use of the intermediate year survey to estimate recruitment does not always provide a good estimate of recruitment (2020 value revised downwards) and that a resampled value (or GM) may in fact provide a better estimate.

The retrospective sensitivity analysis to the exclusion of the age 1 and age 2 catch in the final year also suggests that the assessment is relatively insensitive to the removal of these data in previous years (Figure 3.37) when compared to the assessment runs including all catch data (Intermediate year survey excluded from these assessment runs).
The fits of the model to observations (catch and survey indices on a log scale) are shown in Figures 3.38 to 3.43 . The fits to the survey data appear better at younger ages while the model appears to follow the catch data better at ages 2 and above (age 1 observations are likely to be noisier due to uncertain discard estimates).

The standardised one step ahead residuals are shown in Figure 3.44. There are no major outliers in the residuals, with most lying within $\pm 2$. There are a few patterns apparent in the (discontinued) survey residuals which are rather similar to those observed in previous TSA assessments (ICES, 2019a \& b) and at WKDEM (ICES, 2020): most notably some evidence of a tendency to more positive residuals in the latter half of the WCIBTS.Q1 (at age 1) and WCIBTS.Q4 (at age 2) and some year effects in most of the surveys (years with mostly positive or mostly negative residuals).

The model runs which leave out each survey index in turn are shown in Figure 3.45. With the exception of the period when total catches are excluded from the assessment (catch-scaling factor estimated for 1995-2006), the estimates of SSB and recruitment are relatively robust to the exclusion of the different survey series. Excluding the early Scottish Q4 survey (WCIBTS.Q4) results in higher estimates of SSB, recruitment and catch than the baseline run during this period (when catches area excluded) and excluding the early Scottish Q1 survey much lower estimates. When the WCIBTS.Q4 is excluded, estimates of mean F are lower than the baseline during the first part of this period (to 2000) and higher than the baseline after 2000 while excluding the WCIBTS.Q1 shows the opposite effect. The relative magnitude of the changes when each of these surveys are excluded suggests the WCIBTS.Q1 to be much more influential in the overall assessment of stock trends.

When the SCO.Q4 survey series is excluded there is a downward revision in the estimate of fishing mortality in the final year (although still within the confidence bounds of the estimate) while excluding either the SCO.Q1 or the Irish survey index appears to have little impact on the assessment results.

The retrospective analysis is shown in Figure 3.46. Although the Mohn's rho value for $F$ is within the bounds advised by WKFORBIAS (ICES, 2020c), two of the peels lie outside the confidence intervals of the final assessment run. There appears to be some tendency to over-estimate F. The estimates of mean F appear to be substantially more noisy than SSB. The Mohn's rho values (as $\%$ ) are as follows:

| SSB | Mean F | Recruitment |  |
| :--- | :---: | :--- | :--- |
| -13.4 | 15.4 | 7.5 |  |

In contrast to previous assessments, the recruitment and SSB Mohn's rho do not include the intermediate year in each assessment peel (as this year's assessment does not include an
intermediate year survey and hence SSB and $R$ estimates are not available for this year from the assessment). The Mohn's rho in recruitment is therefore much lower than in previous years as the intermediate year estimate typically shows substantial revisions with the inclusion of additional years' data.

Final parameter estimates from the SAM run are given in Table 3.13. Table 3.14 gives the SAM population numbers-at-age and Table 3.15 the estimated F at-age. A full summary output is given in Table 3.16 (including model estimates of catch and catch scaling parameters).

## Stock status

The summary plot including reference points is shown in Figure 3.47 and the stock-recruitment estimates are shown in Figure 3.48. The estimated SSB shows a steady downward trend until 2006, an increase to 2016 and then a further decline since then. Recruitment has been very low since 2001 and is extremely poor in 2016-2018 and also in 2021. Although fishing mortality declined between 2009 and 2016 to below Flim, it has shown a slight increase since then and is estimated to be just above $\mathrm{F}_{\mathrm{lim}}$ in 2020. It is not known whether, and to what extent, this increase is associated with the discontinuation of the days-at-sea regulation in 2017, which was part of the cod recovery plan.
Estimated SSB in the final year is well below $B_{\lim }$ ( $=14376$ tonnes). Mean F is well above FMSY and has been fluctuating around Flim since 2013. Although the latest assessment shows a flattening off of F since 2013, there has been a clear decrease in mean F since 2009. The decline in mean F is proportionately similar $(\sim 50 \%)$ to the decline in STECF effort (large and small mesh demer$\mathrm{sal} /$ crustacean trawl from both regulated and unregulated fleets), although the mean F does not start to decline until several years after the effort.

### 3.5 Short-term stock projections

Forecasting in SAM takes the form of short-term stochastic projections. A total of 10000 samples are generated from the estimated distribution of survivors. These replicates are then simulated forward according to model and forecast assumptions (see below), using the usual exponential decay equations, but also incorporating the stochastic survival process (using the estimated survival standard deviation) and subject to different catch-options scenarios.
Some modification to the forecast assumptions has been necessary due to the data issues outlined above (lack of intermediate year survey in 2022 and unreliable catch data at age 1 and 2 ). Recruitment in the forecast has been resampled from the assessment estimates for 2016 to 2021. This choice was made due to an apparent further reduction in the level of recruitment in this period (usually a ten year window is chosen). The lack of an intermediate year (2022) recruitment estimate from the assessment (lack of intermediate year survey data) has meant a necessary change to the recruitment assumptions with the resampled recruitment also used for the intermediate year in this year's forecast.

Fishing mortality in the intermediate year (2022) was taken as a three-year average over 2019 to 2021 as an estimate of F status quo (given that there is no particular trend in mean F).

Cod in Division 6a has been fully under the landings obligation since 2019 when a bycatch TAC of 1735 t was set to allow mixed fisheries with a cod bycatch to continue (in contrast to a 0 t TAC with $1.5 \%$ bycatch regulation in previous years). For 2020 and 2021, the bycatch TAC was reduced to 1279 t . These increases in TAC (and the introduction of the LO) appear to have resulted in a significant change in discarding practices since 2019. The partition of catch into landing/discards components in the forecast is therefore based on a recent three-year average (2019-2021)
with the exception of ages 1 and 2 for which the 2021 data are excluded due to the concerns regarding biases in the data for these ages due to lack of OTB_CRU discard samples. A similar approach is also taken for the derivation of forecast mean weights-at-age due to the likely biases in catch mean weights-at-age 1 and 2 in 2021 (See Section 3.3). A summary of the forecast assumptions is given in Table 3.17.
Under the forecast assumption of status quo F, landings in 2022 are predicted to be 1333 t and discards to be 540 t . The SSB in 2023 is forecast to be 2923 t which is well below B lim. This value (2923 t) is similar to that forecast for 2023 from the assessment carried out in 2021 ( 3038 t ) under fishing at $F_{\lim }(=0.73)$, similar to this year's 2022 intermediate year assumption.
The forecast under different catch scenarios for 2023 is shown in Table 3.18. Note that the values that appear in the catch scenarios are medians from the distributions that result from the stochastic forecast.
The forecast stock trajectory under the proposed advice for 2023 (shows an increase in SSB in 2024 (Figure 3.49). Figure 3.50 shows the contribution by recruitment year to SSB in 2024 and catch in 2023 (when fished at FMsY). The assumption regarding recruitment in 2022 to 2024 contribute approximately $50 \%$ of SSB in 2024 and $15 \%$ of the 2023 catch. (Figure 3.50). These values are substantially higher than those reported last year due to the intermediate year recruitment being assumed (rather than an assessment estimate) in this year's forecast.

### 3.5.1 Reference points

Both MSY and precautionary reference points were reconsidered at WKDEM in February 2020 in accordance with ICES guidelines and are shown below (weights in tonnes). The estimate of Fmsy is derived from simulation based on segmented regression stock-recruitment only as both the Ricker and Beverton-Holt stock-recruitment relationships suggest peaks well outside the range of observed values. As in the estimates derived at IBPCOD.6A, yield is defined as catch above MCRS (estimated by assuming a historical discard rate prior to highgrading).

|  | WKMSYREF4 | IBPCod.6a | $\begin{aligned} & \text { WKDEM } \\ & 2020 \end{aligned}$ | Rationale (WKDEM; ICES 2020a) |
| :---: | :---: | :---: | :---: | :---: |
| Blim | 14000 | 14000 | 14376 | Tonnes; SSB consistent with high probability of above average recruitment (SSB in 1992 as estimated by WKDEM) |
| $\mathrm{B}_{\mathrm{pa}}$ | 20000 | 20000 | 20126 | Tonnes; $1.4 \times \mathrm{Bl}_{\text {lim }}$ |
| $F_{\text {lim }}$ | 0.82 | 0.77 | 0.73 | F with $50 \%$ probability of $\mathrm{SSB}<\mathrm{Bl}_{\text {lim }}$ |
| $\mathrm{F}_{\mathrm{pa}}$ | 0.59 | 0.55 | $0.57 \wedge$ | Fp.05; the F that leads to $\mathrm{SSB} \geq \mathrm{B}_{\mathrm{lim}}$ with $95 \%$ probability with ICES AR^ |
| $\mathrm{F}_{\text {MSY }}$ | 0.167 | 0.29 | 0.30 | Based on simulation using a segmented regression stock-recruitment relationship (EqSim) |
| MSY $\mathrm{B}_{\text {trigger }}$ | 20000 | 20000 | 20126 | $\mathrm{B}_{\mathrm{pa}}$ |
| $\mathrm{F}_{\text {MSY }}$ upper | 0.254 | 0.41 | 0.49 | F at 95\% MSY (above $\mathrm{F}_{\text {MSY }}$ ) |
| $\mathrm{F}_{\text {MSY }}$ lower | 0.108 | 0.20 | 0.18 | F at 95 \% MSY (below $\mathrm{F}_{\text {MSY }}$ ) |

^ Updated at WGCSE 2021 following guidance issued by ACOM. Fp. 05 value derived at WKDEM 2020.

### 3.5.2 Management plans

Technical measures applicable to the West of Scotland, including those associated with the cod recovery plan in force up to 2008 (Council Regulation No. 423/2004), the cod long-term management plan in force from 2009 (Council Regulation No. 1342/2008) were amended by Council Regulation No. 1243/2012. The management plan was further amended in 2016 by Council Regulation (EU) 2016/2094 to cover the transitional period in which preparations are ongoing towards multiannual plans for multispecies fisheries. In 2018 the cod management plan was discontinued. Cod in Division 6.a is not included as a named target species in the multiannual plan for Western Waters i.e. only considered as a bycatch species (Council Regulation (EU) 2019/472).

### 3.6 Quality of the assessment

Figure 3.51 shows a comparison between this year's and previous year's assessments. The revised estimates of recruitment and SSB compared to pre-2020 assessments are largely the result of the inclusion of the updated historical catch data at WKDEM (ICES, 2020a). The benchmark changes to the assessment had only minor impact on the perception of the stock.

## Landings

Since the early 1990s the most significant problem with the assessment of this stock is with commercial data. Incorrect reporting of landings, species, quantity and management area, is known to have occurred. Scottish landings (from 2006) are adjusted to include estimates of misreporting (in an attempt to reduce bias in the assessment) and in the five years, 2014-2018, area misreported landings accounted for over $50 \%$ of the total landings although that has reduced to around $20 \%$ in more recent years (and $<5 \%$ in 2021). The misreporting estimates for 2019-2021have been provided by Marine Scotland Compliance based on intelligence and consideration of VMS data (i.e. vessel activity) due to a lack of access to UK VMS data for these years (See Sections 3.1 and 3.3). Estimates for earlier years are derived from VMS data analysis conducted at WKDEM (ICES, 2020a) and these are somewhat lower than MS-C estimates for those years.

## Discards

Although discards have reduced significantly in recent years due to the availability of a bycatch TAC and the implementation of the LO, over the last three years discarding accounts for around $20 \%$ of the total catch. Despite an increase in sampling levels, discard estimates are still very uncertain (approximate $C V=50 \%$ for Scottish large mesh demersal fleet in 2017) contributing to uncertainty in the estimates of mean F.

In 2020 and 2021, discard sampling, and to a lesser extent landings sampling, has been disrupted due to the COVID-19 pandemic, with the most significant impact on the number of samples and seasonal coverage of discard samples from the Nephrops trawl fleet. Due to the lack of Nephrops fishery discard samples in 2021, total discards and catch numbers-at-age 1 and 2 are considered to be underestimated and not included in the assessment. This is likely to result in increased uncertainty in the estimates of recruitment in 2021.

## Biological factors

Cod consumption by seals (derived from diet composition studies and seal abundance estimates) is estimated to be 7632 tonnes ( $95 \%$ CI: 3542-13 937) in 2010 (Hammond and Wilson, 2016)
compared to a TSB estimate of just under 6000 tonnes from the SAM assessment and it has been suggested that seals may be impairing the recovery of this stock. However, there is uncertainty as to whether the seals are actually exploiting the same population as the fishery. Seal foraging mostly occurs on the continental shelf (Russell et al., 2017) including rocky areas which are unsuitable for trawl fishing and are not surveyed on RV trips, while most of the cod landings are taken along the continental shelf edge in the north of Division 6a (STECF, 2016) and thus the seals and fishery are largely operating in different areas. Given the complex stock structure and the presence of coastal cod populations, it is clear there is potential for the seals and fishery to be exploiting different substocks.
The final SAM assessment assumes natural mortality to be a function of stock weight-at-age (Lorenzen, 1996) which are in turn derived from smoothed catch weights-at-age. Natural mortality clearly remains a major source of uncertainty in this assessment and incorrect assumptions regarding its trend and magnitude can have a significant impact on estimates of stock status.

## Stock structure

Stock structure is complex and a number of different subpopulations are known to occur within this area (WK6aCodID; ICES, 2022). The stock assessment therefore represents an assessment of multiple substocks with the northern component accounting for most of the landings since the mid-2000s. The survey distribution plots show that there is an almost complete absence of cod on the shelf in Division 6.a with the majority of the landings and stock concentrated in an area in the north of the region (around the closed area) bordering Division 4.a. A process is underway within ICES to develop a spatial assessment for Northern Shelf cod (North Sea plus Division 6a) which accounts more appropriately for this substock structure.

## Assessment method

The benchmark agreed on the final SAM model configuration by comparing model residuals, AIC and retrospective patterns. There remain some patterns in the residuals particularly in the later surveys which are very noisy and the various sensitivity analyses conducted at WKDEM had little impact on these. Other assessment models also show similar problems. The retrospective analysis in the SAM shows overestimation of fishing mortality during the initial years of decline in mean F (although not persistent across all years of the retrospective analysis), which may suggest the model reacts slowly to changes in fishing mortality.
The input data for this cod assessment are particularly uncertain (both survey indices and commercial data) and as a result, the data can be interpreted in different ways by different assessment methods. The assessment presented by Cook (2019) and a number of exploratory assessments presented at WKDEM show a stock which by 2016 had recovered to levels consistent with those of the 1990s (although with a subsequent decline since then) while the SAM assessment shows little sign of SSB recovery. In this respect, the SAM assessment is very similar to the previous TSA and exploratory a4a assessments considered at the benchmark (ICES, 2020a). The key differences between the Cook (2019) model and the ICES assessment appears to be in the estimates of fishery selectivity and survey catchability and these result in substantial differences in stock trends. An extensive discussion on the plausibility of the estimates can be found in Section 4.3 of ICES (2020a).

Given these model uncertainties, estimates of uncertainty from the final SAM assessment are therefore unlikely to adequately reflect the true uncertainty in the estimates of stock biomass and fishing mortality for this stock.

### 3.6.1 Recommendation for next Benchmark

| problem | solution | expertise neces- <br> sary | suggested time |
| :--- | :--- | :--- | :--- |
| Stock identity - multiple <br> substocks within 6a and <br> linkage with northern <br> North Sea | Evaluate a possible merge be- <br> tween northern North Sea and <br> 6.a cod stocks. Or as an alterna- <br> tive, split area 6.a in two areas <br> North and South. | Scientists from <br>  <br> DTUAqua | Next benchmark although would <br> need collaboration with WGNSSK. |
|  | Requires development of spatial <br> SAM (or alternative) plus deriva- <br> tion of appropriate substock <br> data sets as necessary (catch, <br> survey \& biological) | Explore modelled indices using <br> e.g. delta-logN approach and <br> also modelled ALKs. | Scientists from <br> MSS |
| Noisy survey data | Ahead of next benchmark. |  |  |

${ }^{1}$ MSS = Marine Scotland Science .

### 3.6.2 Management considerations

The fisheries for cod have been fully under the landing obligation from 2019 onwards. In the past they have been managed by a combination of landings limits, area closures and technical measures. The measures taken thus far have not recovered the stock. Although fishing mortality declined between 2009 and 2016, it has shown an increase since then. It is not known whether, and to what extent, this increase is associated with the discontinuation of the days-at-sea regulation in 2017, which was part of the cod recovery plan.

Cod are known to form aggregations, so it is still possible to find areas of high cod density at low stock abundance (as apparent in the Scottish Q1 survey in particular). This can lead to high caches in localized areas, generating high fishing mortality even with low fishing effort. The impact of this could potentially be reduced by the use of temporary spatial closures.

The fishing opportunities regulation explicitly made the stock a bycatch species from 2012 to 2018. Allowing landings up to $1.5 \%$ of the live weight of the total catch can cause a perverse incentive for vessels to increase catches of other species and does not inhibit the catch of cod.

Although the UK 'Buyers and Sellers' and Irish 'Sales Notes' legislation is considered to have reduced underreporting from 2006, discard data showed increased discards at-ages one and two and a change in discard practices such that fish are discarded at older ages from 2006-2018 (i.e. such that the discards were largely highgrading). With the full implementation of the landing obligation in 2019 for fisheries catching cod, a bycatch TAC of 1735 t was set to allow mixed
fisheries with a bycatch of cod to continue. The fishery has responded to this by reducing discards, particularly at older ages. The forecast assumes that this discarding behaviour will continue in future. The bycatch TAC has been reduced to 1279 t for 2020 and 2021.

Estimates of area misreporting (landings believed to be taken in Division 6.a and reported elsewhere) imply ICES landings estimates that are in excess of TAC. Area misreported landings accounted for around $20 \%$ of the total landings in 2019-2021 which is a reduction on previous years.

Cod is taken in mixed demersal fisheries, and in Division 6.a is a bycatch species. To greatly reduce cod catch would likely result in having to greatly reduce harvesting of other stocks such as haddock, whiting and anglerfish. It is also important the bycatch from the Nephrops fleet is closely monitored (including discard observations). Typically, large trawl gear vessels targeting finfish are responsible for around $90 \%$ of cod catches in Division 6.a, the Nephrops fleet take approximately $4 \%$ and the remainder are taken by other gears, including longliners and gillnets. (Note that data for 2021 are unreliable due to lack of OTB_CRU discard sampling).
A report by the Sea Mammal Research unit (Hammond and Harris, 2006) gives estimates of cod consumed by grey seals to the west of Scotland. Although highly uncertain, the estimates suggest predation mortality on cod is significant and this may impair the ability of the cod stock to recover, but data are limited (Cook et al., 2015).

Cod to the west of Scotland (6a.) are believed to comprise of at least two subpopulations and potentially linked to cod in the North Sea (4a). The current assessments and management do not capture this dynamic as they are treated independently.

### 3.7 References

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Table 3.1. Cod.27.6a. ICES official catch statistics.

| Country |  |  |  | シ 픈 픈․ |  |  | $\begin{aligned} & \text { ס } \\ & \text { 들 } \\ & \underline{N D} \end{aligned}$ |  |  | $\begin{aligned} & \stackrel{.}{\overline{0}} \\ & \stackrel{0}{n} \end{aligned}$ |  |  | ソ | - 즌 | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 48 | - | - | 7411 | 66 | - | 2564 | - | 204 | 28 | 260 | 8032 | - |  | 18613 |
| 1986 | 88 | - | - | 5096 | 53 | - | 1704 | - | 174 | - | 160 | 4251 | - |  | 11526 |
| 1987 | 33 | 4 | - | 5044 | 12 | - | 2442 | - | 77 | - | 444 | 11143 | - |  | 19199 |
| 1988 | 44 | 1 | 11 | 7669 | 25 | - | 2551 | - | 186 | - | 230 | 8465 | - |  | 19182 |
| 1989 | 28 | 3 | 26 | 3640 | 281 | - | 1642 | - | 207 | 85 | 278 | 9236 | - |  | 15426 |
| 1990 | - | 2 | - | 2220 | 586 | - | 1200 | - | 150 | - | 230 | 7389 | - |  | 11777 |
| 1991 | 6 | 2 | - | 2503 | 60 | - | 761 | - | 40 | - | 511 | 6751 | - |  | 10634 |
| 1992 | - | 3 | - | 1957 | 5 | - | 761 | - | 171 | - | 577 | 5543 | - |  | 9017 |
| 1993 | 22 | 2 | - | 3047 | 94 | - | 645 | - | 72 | - | 524 | 6069 | - |  | 10475 |
| 1994 | 1 | + | - | 2488 | 100 | - | 825 | - | 51 | - | 419 | 5247 | - |  | 9131 |
| 1995 | 2 | 4 | - | 2533 | 18 | - | 1054 | - | 61 | 16 | 450 | 5522 | - |  | 9660 |
| 1996 | + | 2 | - | 2253 | 63 | - | 1286 | - | 137 | + | 457 | 5382 | - |  | 9580 |
| 1997 | 11 | - | - | 956 | 5 | - | 708 | 2 | 36 | 6 | 779 | 4489 | - |  | 6992 |
| 1998 | 1 | - | - | 714 | 6 | - | 478 | 1 | 36 | 42 | 474 | 3919 | - |  | 5671 |
| 1999 | + | + | - | 842 | 8 | - | 223 | - | 79 | 45 | 381 | 2711 | - |  | 4289 |
| 2000 | + | - | - | 236 | 6 | - | 357 | - | 114 | 14 | 280 | 2057 | - |  | 3064 |
| 2001 | 2 | - | - | 391 | 4 | - | 319 | - | 39 | 3 | 138 | 1544 | - |  | 2440 |
| 2002 | + | - | - | 208 | + | - | 210 | - | 88 | 11 | 195 | 1519 | - |  | 2231 |
| 2003 | - | - | - | 172 | + | - | 120 | - | 45 | 3 | 79 | 879 | - |  | 1298 |
| 2004 | - | - | 2 | 91 | - | - | 34 | - | 10 | - | 46 | 413 | - |  | 596 |
| 2005 | - | - | - | 107 | - | - | 28 | - | 17 | - | 25 | 243 | - |  | 420 |
| 2006 | - | - | 1 | 108 | 2 | - | 18 | - | 30 | - | 14 | 318 | - |  | 491 |
| 2007 | - | - | 12 | 92 | 2 | - | 70 | - | 30 | - | 21 | 260 | - |  | 487 |
| 2008 | - | - | 1 | 82 | 1 | - | 58 | - | 65 | - | 6 | 232 | - |  | 445 |
| 2009 | - | - |  | 74 | - | - | 24 | - | 18 | - | 14 | 104 | - |  | 234 |
| 2010 | - | - | - | 60 | - | - | 49 | - | 21 | - | 4 | 115 | - |  | 249 |


| Country |  |  |  |  |  |  |  |  |  | $\begin{aligned} & \cdot \frac{\cong}{0} \\ & \text { in } \end{aligned}$ |  |  | $\underset{ }{〕}$ | 倞 $\sum_{\infty}^{n}$ | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | - | - | - | 49 | - | - | 41 | - | 8 | - | 3 | 107 | - |  | 208 |
| 2012 | - | - | - | 4 | - | - | 18 | - | 2 | - | 2 | 135 | - |  | 161 |
| 2013 | - | - | - | 3 | - | - | 14 | - | 24 | - | 1 | 130 | - |  | 172 |
| 2014 | - | - | - | 5 | - | - | 12 | - | 13 | - | 9 | 121 | - |  | 160 |
| 2015 | - | - | - | 11 | - | - | 17 | - | 59 | - | - | - | 168 |  | 256 |
| 2016 | - | 11 | - | 86 | - | 1 | 28 | - | 39 | - | - | - | 183 |  | 348 |
| 2017 | - | 1 | - | 119 | - | - | 19 | - | 14 | - | - | - | 200 |  | 352 |
| 2018 | - | + | + | 101 | - | - | 12 | - | 37 | - | - | - | 217 |  | 367 |
| 2019 | - | - | - | 142 | - | - | $\wedge$ | - | 47 | 31 | - | - | 1224 | + | 1443^ |
| 2020* | - | - | - | 139 | - | 3 | 65 | - | 4 | 32 | - | - | 738 | 2 | 983 |
| 2021* | - | - | - | 162 | - | - | 98 | - | - | 27 | - | - | 923 | + | 1209 |

* Preliminary.
$+<0.5$ tonnes.
${ }^{\wedge}$ Incomplete/missing due to part of the data being unavailable under national GDPR clauses.

Table 3.2. Cod.27.6a. Landings (reported into 6a and area misreported), discards, BMS and catch (tonnes) estimates, as used by the WG (caton from InterCatch).

| Year | Landings |  | Discards | BMS | Catch |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | reported | misreported |  |  |  |
| 1981 | 23865 |  | 303 |  | 24168 |
| 1982 | 21511 |  | 571 |  | 22082 |
| 1983 | 21305 |  | 197 |  | 21503 |
| 1984 | 21272 |  | 329 |  | 21601 |
| 1985 | 18607 |  | 963 |  | 19570 |
| 1986 | 11820 |  | 263 |  | 12083 |
| 1987 | 18971 |  | 2388 |  | 21358 |
| 1988 | 20413 |  | 368 |  | 20781 |
| 1989 | 17169 |  | 2076 |  | 19246 |
| 1990 | 12175 |  | 571 |  | 12746 |
| 1991 | 10927 |  | 622 |  | 11549 |
| 1992 | 9086 |  | 1779 |  | 10865 |
| 1993 | 10314 |  | 139 |  | 10453 |
| 1994 | 8928 |  | 661 |  | 9588 |
| 1995 | 9439 |  | 141 |  | 9580 |
| 1996 | 9427 |  | 63 |  | 9489 |
| 1997 | 7034 |  | 499 |  | 7533 |
| 1998 | 5714 |  | 538 |  | 6252 |
| 1999 | 4201 |  | 69 |  | 4270 |
| 2000 | 2977 |  | 821 |  | 3798 |
| 2001 | 2347 |  | 92 |  | 2439 |
| 2002 | 2243 |  | 480 |  | 2722 |
| 2003 | 1292 |  | 60 |  | 1353 |
| 2004 | 573 |  | 78 |  | 651 |
| 2005 | 516 |  | 54 |  | 570 |
| 2006 | 470 | 34 | 461 |  | 965 |


| Year | Landings |  | Discards | BMS | Catch |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | reported | misreported |  |  |  |
| 2007 | 485 | 30 | 1651 |  | 2166 |
| 2008 | 460 | 102 | 1037 |  | 1598 |
| 2009 | 231 | 54 | 1287 |  | 1572 |
| 2010 | 239 | 119 | 1575 |  | 1933 |
| 2011 | 211 | 130 | 3867 |  | 4208 |
| 2012 | 162 | 65 | 1914 |  | 2141 |
| 2013 | 172 | 93 | 1870 |  | 2136 |
| 2014 | 161 | 234 | 3369 |  | 3764 |
| 2015 | 258 | 270 | 2498 |  | 3026 |
| 2016 | 336 | 272 | 1499 |  | 2108 |
| 2017 | 355 | 320 | 3519 |  | 4195 |
| 2018 | 378 | 613 | 2429 |  | 3419 |
| 2019 | 1489 | 571 | 204 |  | 2264 |
| 2020 | 941 | 332 | 307 | 2.5 | 1583 |
| 2021 | 1215 | 49 | 642 | + | 1907 |

$+<0.5$ tonnes.

Table 3.3. Cod.27.6a. Landings-at-age (thousands). Values for 2006 onwards include an adjustment for area misreporting.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 461 | 7016 | 3220 | 904 | 182 | 29 | 20 |
| 1982 | 1827 | 1673 | 3206 | 1189 | 367 | 111 | 33 |
| 1983 | 2335 | 4515 | 1118 | 1400 | 468 | 148 | 60 |
| 1984 | 2143 | 2360 | 2564 | 448 | 555 | 185 | 59 |
| 1985 | 1355 | 5069 | 1269 | 1091 | 140 | 167 | 79 |
| 1986 | 792 | 1486 | 2055 | 411 | 191 | 40 | 30 |
| 1987 | 7873 | 4837 | 988 | 905 | 137 | 56 | 26 |
| 1988 | 1008 | 8336 | 2193 | 278 | 210 | 39 | 20 |
| 1989 | 2017 | 1082 | 3858 | 709 | 113 | 69 | 33 |
| 1990 | 513 | 4024 | 432 | 924 | 170 | 23 | 11 |
| 1991 | 1518 | 1728 | 1805 | 188 | 266 | 70 | 23 |
| 1992 | 1407 | 1868 | 575 | 720 | 69 | 58 | 24 |
| 1993 | 328 | 3596 | 1050 | 131 | 183 | 24 | 36 |
| 1994 | 942 | 1207 | 1545 | 280 | 56 | 51 | 20 |
| 1995 | 753 | 2750 | 700 | 630 | 70 | 15 | 11 |
| 1996 | 341 | 2331 | 1210 | 247 | 204 | 31 | 13 |
| 1997 | 1414 | 1067 | 989 | 281 | 66 | 62 | 7 |
| 1998 | 310 | 3318 | 293 | 174 | 57 | 16 | 9 |
| 1999 | 132 | 884 | 1047 | 64 | 48 | 24 | 9 |
| 2000 | 765 | 532 | 211 | 231 | 15 | 12 | 13 |
| 2001 | 96 | 1241 | 155 | 63 | 52 | 3 | 4 |
| 2002 | 337 | 340 | 522 | 41 | 13 | 14 | 4 |
| 2003 | 53 | 487 | 93 | 120 | 7 | 2 | 2 |
| 2004 | 45 | 99 | 90 | 12 | 27 | 3 | 1 |
| 2005 | 37 | 124 | 46 | 40 | 7 | 6 | 0 |
| 2006 | 18 | 97 | 78 | 23 | 14 | 2 | 1 |
| 2007 | 7 | 170 | 53 | 28 | 2 | 3 | 2 |
| 2008 | 0 | 20 | 106 | 21 | 13 | 1 | 2 |
| 2009 | 1 | 9 | 10 | 40 | 6 | 1 | 0 |
| 2010 | 6 | 80 | 26 | 20 | 11 | 1 | 1 |
| 2011 | 0 | 29 | 51 | 18 | 4 | 6 | 1 |
| 2012 | 1 | 1 | 18 | 24 | 3 | 2 | 2 |
| 2013 | 0 | 8 | 7 | 39 | 9 | 2 | 1 |
| 2014 | 0 | 5 | 73 | 34 | 25 | 2 | 0 |
| 2015 | 0 | 44 | 40 | 29 | 21 | 19 | 1 |
| 2016 | 1 | 17 | 82 | 52 | 17 | 9 | 11 |


|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2017 | 0 | 13 | 52 | 47 | 46 | 13 | 3 |
| 2018 | 2 | 10 | 28 | 78 | 51 | 32 | 11 |
| 2019 | 9 | 21 | 129 | 89 | 142 | 57 | 13 |
| 2020 | 7 | 75 | 9 | 55 | 44 | 53 | 30 |
| 2021 | 1 | 29 | 49 | 47 | 8 | 12 |  |

Table 3.4. Cod.27.6a. Mean weight-at-age in landings (kg).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.55 | 1.166 | 2.839 | 4.923 | 7.518 | 9.314 | 10.328 |
| 1982 | 0.692 | 1.468 | 2.737 | 4.749 | 6.113 | 7.227 | 9.856 |
| 1983 | 0.583 | 1.265 | 2.995 | 4.398 | 6.305 | 8.084 | 9.744 |
| 1984 | 0.735 | 1.402 | 3.168 | 5.375 | 6.601 | 8.606 | 10.35 |
| 1985 | 0.628 | 1.183 | 2.597 | 4.892 | 6.872 | 8.344 | 9.766 |
| 1986 | 0.71 | 1.211 | 2.785 | 4.655 | 6.336 | 8.283 | 9.441 |
| 1987 | 0.531 | 1.312 | 2.783 | 4.574 | 6.161 | 7.989 | 10.062 |
| 1988 | 0.806 | 1.182 | 2.886 | 5.145 | 6.993 | 8.204 | 9.803 |
| 1989 | 0.704 | 1.298 | 2.425 | 4.737 | 7.027 | 7.52 | 9.594 |
| 1990 | 0.613 | 1.275 | 2.815 | 4.314 | 7.021 | 9.027 | 11.671 |
| 1991 | 0.64 | 1.095 | 2.618 | 4.346 | 6.475 | 8.134 | 10.076 |
| 1992 | 0.686 | 1.293 | 2.607 | 4.268 | 6.19 | 7.844 | 10.598 |
| 1993 | 0.775 | 1.316 | 2.94 | 4.646 | 6.244 | 7.802 | 8.409 |
| 1994 | 0.644 | 1.292 | 2.899 | 4.71 | 6.389 | 8.423 | 8.409 |
| 1995 | 0.606 | 1.148 | 2.857 | 4.956 | 6.771 | 8.539 | 9.505 |
| 1996 | 0.667 | 1.221 | 2.738 | 5.056 | 6.892 | 8.088 | 10.759 |
| 1997 | 0.595 | 1.21 | 2.571 | 4.805 | 6.952 | 7.821 | 9.63 |
| 1998 | 0.605 | 1.061 | 2.264 | 4.506 | 6.104 | 8.017 | 9.612 |
| 1999 | 0.691 | 1.039 | 2.194 | 4.688 | 6.486 | 8.252 | 9.439 |
| 2000 | 0.689 | 1.261 | 2.457 | 4.126 | 6.666 | 7.917 | 8.392 |
| 2001 | 0.654 | 0.988 | 2.679 | 4.568 | 5.86 | 7.741 | 9.386 |
| 2002 | 0.668 | 1.14 | 2.33 | 4.841 | 6.175 | 7.192 | 9.548 |
| 2003 | 0.659 | 1.046 | 2.272 | 3.82 | 5.932 | 8.022 | 8.681 |
| 2004 | 0.605 | 1.026 | 2.191 | 4.398 | 6.033 | 8.242 | 9.84 |
| 2005 | 0.75 | 1.109 | 2.425 | 3.969 | 4.775 | 6.616 | 10.214 |
| 2006 | 0.659 | 1.176 | 2.239 | 3.813 | 6.16 | 7.759 | 11.041 |
| 2007 | 0.728 | 1.127 | 2.592 | 4.322 | 6.503 | 7.738 | 8.83 |
| 2008 | 0.556 | 1.157 | 3.067 | 4.843 | 6.283 | 7.964 | 8.487 |
| 2009 | 0.974 | 2.038 | 2.861 | 4.781 | 6.004 | 8.327 | 9.137 |
| 2010 | 0.936 | 1.468 | 2.918 | 4.064 | 5.785 | 9.158 | 10.275 |
| 2011 | 0 | 1.804 | 2.811 | 4.51 | 5.842 | 6.528 | 9.837 |
| 2012 | 0.661 | 1.797 | 3.118 | 5.331 | 6.428 | 7.617 | 8.695 |
| 2013 | 0.957 | 1.368 | 2.933 | 4.075 | 6.135 | 7.144 | 9.842 |
| 2014 | 1.028 | 1.6 | 2.097 | 3.051 | 4.693 | 5.503 | 7.207 |
| 2015 | 0.914 | 2.406 | 2.958 | 3.844 | 5.455 | 5.558 | 9.158 |
| 2016 | 0.713 | 1.429 | 2.367 | 3.917 | 5.137 | 6.596 | 7.622 |
| 2017 | 0.902 | 1.229 | 2.063 | 4.533 | 5.616 | 5.081 | 9.243 |


|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2018 | 0.871 | 1.686 | 2.761 | 4.163 | 5.427 | 6.427 | 8.575 |
| 2019 | 0.857 | 1.159 | 2.962 | 4.242 | 5.461 | 7.045 | 8.841 |
| 2020 | 0.618 | 1.310 | 2.308 | 4.763 | 5.957 | 6.362 | 6.448 |
| 2021 | 0.908 | 1.207 | 2.760 | 3.518 | 5.443 | 7.316 | 7.377 |

Table 3.5. Cod.27.6a. Discard numbers-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 54 | 907 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 1808 | 8 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 843 | 25 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 1088 | 11 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 5188 | 114 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 970 | 14 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 14358 | 12 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 231 | 1059 | 2 | 0 | 0 | 0 | 0 |
| 1989 | 6243 | 6 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 4181 | 41 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 2518 | 14 | 2 | 0 | 0 | 0 | 0 |
| 1992 | 7385 | 143 | 3 | 0 | 0 | 0 | 0 |
| 1993 | 279 | 84 | 1 | 0 | 0 | 0 | 0 |
| 1994 | 2743 | 6 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 625 | 56 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 191 | 50 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 1521 | 34 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 790 | 972 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 230 | 5 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 2882 | 33 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 176 | 115 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 1051 | 199 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 124 | 27 | 7 | 0 | 0 | 0 | 0 |
| 2004 | 238 | 23 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 127 | 22 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 1058 | 45 | 25 | 2 | 3 | 1 | 0 |
| 2007 | 283 | 1321 | 46 | 35 | 2 | 3 | 0 |
| 2008 | 64 | 151 | 416 | 3 | 1 | 0 | 0 |
| 2009 | 590 | 157 | 116 | 146 | 8 | 7 | 0 |
| 2010 | 410 | 810 | 150 | 17 | 7 | 0 | 0 |
| 2011 | 303 | 579 | 1255 | 102 | 1 | 4 | 0 |
| 2012 | 1029 | 180 | 605 | 78 | 0 | 0 | 0 |
| 2013 | 2175 | 346 | 220 | 167 | 24 | 0 | 3 |
| 2014 | 913 | 948 | 644 | 116 | 45 | 2 | 0 |
| 2015 | 264 | 571 | 620 | 72 | 18 | 2 | 0 |
| 2016 | 1253 | 377 | 189 | 94 | 13 | 0 | 0 |
| 2017 | 240 | 429 | 912 | 223 | 43 | 5 | 0 |


|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2018 | 87 | 447 | 206 | 300 | 54 | 18 | 6 |
| 2019 | 248 | 112 | 49 | 6 | 1 | 0 | 0 |
| 2020 | 304 | 173 | 16 | 10 | 0 | 0 | 0 |
| 2021 | 6 | 174 | 131 | 1 | 6 | 0 | 0 |

Table 3.6. Cod.27.6a. Mean weight-at-age in discards (kg).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.135 | 0.326 |  |  |  |  |  |
| 1982 | 0.314 | 0.392 |  |  |  |  |  |
| 1983 | 0.223 | 0.374 |  |  |  |  |  |
| 1984 | 0.298 | 0.435 |  |  |  |  |  |
| 1985 | 0.178 | 0.346 |  |  |  |  |  |
| 1986 | 0.267 | 0.305 |  |  |  |  |  |
| 1987 | 0.166 | 0.37 |  |  |  |  |  |
| 1988 | 0.296 | 0.283 |  |  |  |  |  |
| 1989 | 0.332 | 0.59 |  |  |  |  |  |
| 1990 | 0.132 | 0.454 |  |  |  |  |  |
| 1991 | 0.245 | 0.351 |  |  |  |  |  |
| 1992 | 0.22 | 1.03 | 2.382 |  |  |  |  |
| 1993 | 0.239 | 0.812 | 3.723 |  |  |  |  |
| 1994 | 0.24 | 0.365 |  |  |  |  |  |
| 1995 | 0.203 | 0.256 |  |  |  |  |  |
| 1996 | 0.226 | 0.389 |  |  |  |  |  |
| 1997 | 0.321 | 0.328 |  |  |  |  |  |
| 1998 | 0.23 | 0.367 | 0.59 |  |  |  |  |
| 1999 | 0.294 | 0.299 |  |  |  |  |  |
| 2000 | 0.28 | 0.421 |  |  |  |  |  |
| 2001 | 0.248 | 0.417 |  |  |  |  |  |
| 2002 | 0.263 | 1.021 |  |  |  |  |  |
| 2003 | 0.311 | 0.6 | 0.388 |  |  |  |  |
| 2004 | 0.261 | 0.576 |  |  |  |  |  |
| 2005 | 0.242 | 0.483 | 0.803 |  |  |  |  |
| 2006 | 0.276 | 1.346 | 2.786 | 3.501 | 6.242 | 5.581 | 11.151 |
| 2007 | 0.196 | 0.948 | 3.014 | 4.457 | 4.985 | 10.635 |  |
| 2008 | 0.224 | 0.999 | 2.049 | 3.853 | 5.216 |  |  |
| 2009 | 0.264 | 1.333 | 2.296 | 3.834 | 6.051 | 6.985 | 9.119 |
| 2010 | 0.273 | 1.274 | 2.268 | 3.218 | 3.245 |  |  |
| 2011 | 0.266 | 1.072 | 2.213 | 2.993 | 4.891 | 4.168 |  |
| 2012 | 0.142 | 1.118 | 2.179 | 3.222 |  |  |  |
| 2013 | 0.125 | 1.155 | 2.11 | 3.05 | 5.029 |  | 6.269 |
| 2014 | 0.15 | 1.21 | 2.39 | 3.066 | 3.998 | 4.349 |  |
| 2015 | 0.404 | 1.063 | 2.33 | 3.428 | 4.414 | 6.103 |  |
| 2016 | 0.205 | 1.096 | 2.212 | 3.759 | 4.435 |  |  |
| 2017 | 0.262 | 1.048 | 2.183 | 3.473 | 4.397 | 7.714 |  |


|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2018 | 0.217 | 1.046 | 2.219 | 3.649 | 5.3 | 4.98 | 2.117 |
| 2019 | 0.226 | 0.548 | 1.397 | 2.318 | 3.516 |  |  |
| 2020 | 0.167 | 0.922 | 3.199 | 4.763 |  |  |  |
| 2021 | 0.708 | 1.348 | 2.821 | 4.309 | 5.175 |  |  |

Table 3.7. Cod.27.6a. Total catch-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 515 | 7923 | 3220 | 904 | 182 | 29 | 20 |
| 1982 | 3635 | 1681 | 3206 | 1189 | 367 | 111 | 33 |
| 1983 | 3178 | 4540 | 1118 | 1400 | 468 | 148 | 60 |
| 1984 | 3231 | 2371 | 2564 | 448 | 555 | 185 | 59 |
| 1985 | 6543 | 5183 | 1269 | 1091 | 140 | 167 | 79 |
| 1986 | 1762 | 1500 | 2055 | 411 | 191 | 40 | 30 |
| 1987 | 22231 | 4849 | 988 | 905 | 137 | 56 | 26 |
| 1988 | 1239 | 9395 | 2195 | 278 | 210 | 39 | 20 |
| 1989 | 8260 | 1088 | 3858 | 709 | 113 | 69 | 33 |
| 1990 | 4694 | 4065 | 432 | 924 | 170 | 23 | 11 |
| 1991 | 4036 | 1742 | 1807 | 188 | 266 | 70 | 23 |
| 1992 | 8792 | 2011 | 578 | 720 | 69 | 58 | 24 |
| 1993 | 607 | 3680 | 1051 | 131 | 183 | 24 | 36 |
| 1994 | 3685 | 1213 | 1545 | 280 | 56 | 51 | 20 |
| 1995 | 1378 | 2806 | 700 | 630 | 70 | 15 | 11 |
| 1996 | 532 | 2381 | 1210 | 247 | 204 | 31 | 13 |
| 1997 | 2935 | 1101 | 989 | 281 | 66 | 62 | 7 |
| 1998 | 1100 | 4290 | 293 | 174 | 57 | 16 | 9 |
| 1999 | 362 | 889 | 1047 | 64 | 48 | 24 | 9 |
| 2000 | 3647 | 565 | 211 | 231 | 15 | 12 | 13 |
| 2001 | 272 | 1356 | 155 | 63 | 52 | 3 | 4 |
| 2002 | 1388 | 539 | 522 | 41 | 13 | 14 | 4 |
| 2003 | 176 | 514 | 100 | 120 | 7 | 2 | 2 |
| 2004 | 282 | 122 | 90 | 12 | 27 | 3 | 1 |
| 2005 | 163 | 146 | 46 | 40 | 7 | 6 | 0 |
| 2006 | 1076 | 143 | 104 | 25 | 17 | 3 | 1 |
| 2007 | 290 | 1492 | 100 | 64 | 5 | 6 | 2 |
| 2008 | 64 | 171 | 522 | 24 | 15 | 1 | 2 |
| 2009 | 591 | 166 | 126 | 186 | 14 | 8 | 1 |
| 2010 | 416 | 889 | 175 | 37 | 17 | 1 | 1 |
| 2011 | 303 | 608 | 1307 | 120 | 5 | 10 | 1 |
| 2012 | 1030 | 181 | 623 | 101 | 3 | 2 | 2 |
| 2013 | 2175 | 355 | 228 | 206 | 33 | 2 | 4 |
| 2014 | 913 | 953 | 717 | 149 | 70 | 4 | 0 |
| 2015 | 264 | 615 | 660 | 102 | 39 | 21 | 1 |
| 2016 | 1254 | 394 | 271 | 146 | 30 | 9 | 11 |
| 2017 | 240 | 442 | 963 | 270 | 89 | 18 | 3 |


|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2018 | 88 | 457 | 235 | 378 | 105 | 49 | 16 |
| 2019 | 256 | 132 | 178 | 95 | 142 | 57 | 13 |
| 2020 | 311 | 248 | 26 | 65 | 44 | 53 | 30 |
| 2021 | 6 | 203 | 359 | 50 | 53 | 8 | 12 |

Table 3.8. Cod.27.6a. Mean weight-at-age (kg) in total catch.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.506 | 1.07 | 2.839 | 4.923 | 7.518 | 9.314 | 10.328 |
| 1982 | 0.504 | 1.463 | 2.737 | 4.749 | 6.113 | 7.227 | 9.856 |
| 1983 | 0.488 | 1.26 | 2.995 | 4.398 | 6.305 | 8.084 | 9.744 |
| 1984 | 0.588 | 1.398 | 3.168 | 5.375 | 6.601 | 8.606 | 10.35 |
| 1985 | 0.271 | 1.165 | 2.597 | 4.892 | 6.872 | 8.344 | 9.766 |
| 1986 | 0.466 | 1.203 | 2.785 | 4.655 | 6.336 | 8.283 | 9.441 |
| 1987 | 0.295 | 1.31 | 2.783 | 4.574 | 6.161 | 7.989 | 10.062 |
| 1988 | 0.711 | 1.081 | 2.883 | 5.145 | 6.993 | 8.204 | 9.803 |
| 1989 | 0.423 | 1.294 | 2.425 | 4.737 | 7.027 | 7.52 | 9.594 |
| 1990 | 0.185 | 1.267 | 2.815 | 4.314 | 7.021 | 9.027 | 11.671 |
| 1991 | 0.394 | 1.089 | 2.615 | 4.346 | 6.475 | 8.134 | 10.076 |
| 1992 | 0.295 | 1.274 | 2.606 | 4.268 | 6.19 | 7.844 | 10.598 |
| 1993 | 0.529 | 1.304 | 2.941 | 4.646 | 6.244 | 7.802 | 8.409 |
| 1994 | 0.343 | 1.287 | 2.899 | 4.71 | 6.389 | 8.423 | 8.409 |
| 1995 | 0.423 | 1.13 | 2.857 | 4.956 | 6.771 | 8.539 | 9.505 |
| 1996 | 0.509 | 1.204 | 2.738 | 5.056 | 6.892 | 8.088 | 10.759 |
| 1997 | 0.453 | 1.183 | 2.571 | 4.805 | 6.952 | 7.821 | 9.63 |
| 1998 | 0.336 | 0.904 | 2.264 | 4.506 | 6.104 | 8.017 | 9.612 |
| 1999 | 0.439 | 1.035 | 2.194 | 4.688 | 6.486 | 8.252 | 9.439 |
| 2000 | 0.366 | 1.212 | 2.457 | 4.126 | 6.666 | 7.917 | 8.392 |
| 2001 | 0.391 | 0.94 | 2.679 | 4.568 | 5.86 | 7.741 | 9.386 |
| 2002 | 0.361 | 1.096 | 2.33 | 4.841 | 6.175 | 7.192 | 9.548 |
| 2003 | 0.415 | 1.023 | 2.14 | 3.82 | 5.932 | 8.022 | 8.681 |
| 2004 | 0.316 | 0.943 | 2.191 | 4.398 | 6.033 | 8.242 | 9.84 |
| 2005 | 0.356 | 1.014 | 2.425 | 3.969 | 4.775 | 6.616 | 10.214 |
| 2006 | 0.282 | 1.23 | 2.373 | 3.789 | 6.175 | 7.002 | 11.046 |
| 2007 | 0.209 | 0.969 | 2.788 | 4.397 | 5.726 | 9.174 | 8.83 |
| 2008 | 0.224 | 1.018 | 2.256 | 4.715 | 6.189 | 7.964 | 8.487 |
| 2009 | 0.266 | 1.372 | 2.342 | 4.039 | 6.03 | 7.222 | 9.111 |
| 2010 | 0.282 | 1.291 | 2.363 | 3.683 | 4.784 | 9.158 | 10.275 |
| 2011 | 0.266 | 1.107 | 2.237 | 3.221 | 5.722 | 5.507 | 9.837 |
| 2012 | 0.142 | 1.12 | 2.205 | 3.713 | 6.428 | 7.617 | 8.695 |
| 2013 | 0.125 | 1.16 | 2.137 | 3.243 | 5.336 | 7.144 | 7.145 |
| 2014 | 0.15 | 1.212 | 2.36 | 3.063 | 4.245 | 4.984 | 7.207 |
| 2015 | 0.405 | 1.159 | 2.368 | 3.548 | 4.964 | 5.612 | 9.158 |
| 2016 | 0.206 | 1.11 | 2.259 | 3.815 | 4.834 | 6.596 | 7.622 |
| 2017 | 0.263 | 1.053 | 2.177 | 3.656 | 5.032 | 5.746 | 9.243 |


|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2018 | 0.229 | 1.06 | 2.285 | 3.755 | 5.362 | 5.909 | 6.304 |
| 2019 | 0.248 | 0.644 | 2.532 | 4.112 | 5.450 | 7.045 | 8.841 |
| 2020 | 0.178 | 1.039 | 2.873 | 4.763 | 5.957 | 6.362 | 6.448 |
| 2021 | 0.730 | 1.327 | 2.782 | 3.534 | 5.413 | 7.316 | 7.377 |

Table 3.9. Cod.27.6a. Mean weight-at-age (kg) in stock.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.496 | 1.262 | 2.888 | 4.854 | 6.932 | 8.447 | 10.100 |
| 1982 | 0.488 | 1.256 | 2.874 | 4.838 | 6.891 | 8.421 | 10.077 |
| 1983 | 0.480 | 1.250 | 2.859 | 4.821 | 6.849 | 8.394 | 10.054 |
| 1984 | 0.473 | 1.244 | 2.844 | 4.803 | 6.807 | 8.367 | 10.030 |
| 1985 | 0.465 | 1.238 | 2.827 | 4.783 | 6.766 | 8.340 | 10.008 |
| 1986 | 0.457 | 1.232 | 2.810 | 4.761 | 6.724 | 8.314 | 9.985 |
| 1987 | 0.450 | 1.225 | 2.793 | 4.738 | 6.682 | 8.287 | 9.961 |
| 1988 | 0.442 | 1.219 | 2.776 | 4.716 | 6.641 | 8.260 | 9.937 |
| 1989 | 0.434 | 1.213 | 2.760 | 4.696 | 6.599 | 8.232 | 9.912 |
| 1990 | 0.427 | 1.207 | 2.745 | 4.681 | 6.557 | 8.203 | 9.885 |
| 1991 | 0.419 | 1.201 | 2.729 | 4.671 | 6.514 | 8.173 | 9.857 |
| 1992 | 0.411 | 1.195 | 2.712 | 4.666 | 6.472 | 8.142 | 9.827 |
| 1993 | 0.404 | 1.188 | 2.691 | 4.665 | 6.429 | 8.109 | 9.796 |
| 1994 | 0.396 | 1.182 | 2.666 | 4.664 | 6.386 | 8.075 | 9.764 |
| 1995 | 0.388 | 1.176 | 2.636 | 4.660 | 6.342 | 8.038 | 9.731 |
| 1996 | 0.380 | 1.170 | 2.602 | 4.649 | 6.298 | 7.999 | 9.698 |
| 1997 | 0.373 | 1.164 | 2.564 | 4.628 | 6.254 | 7.957 | 9.664 |
| 1998 | 0.365 | 1.158 | 2.526 | 4.596 | 6.210 | 7.913 | 9.628 |
| 1999 | 0.357 | 1.152 | 2.489 | 4.553 | 6.165 | 7.866 | 9.591 |
| 2000 | 0.350 | 1.145 | 2.455 | 4.500 | 6.120 | 7.816 | 9.551 |
| 2001 | 0.342 | 1.139 | 2.426 | 4.440 | 6.075 | 7.764 | 9.509 |
| 2002 | 0.334 | 1.133 | 2.401 | 4.377 | 6.029 | 7.708 | 9.463 |
| 2003 | 0.327 | 1.127 | 2.381 | 4.310 | 5.983 | 7.650 | 9.412 |
| 2004 | 0.319 | 1.121 | 2.364 | 4.242 | 5.937 | 7.588 | 9.356 |
| 2005 | 0.311 | 1.115 | 2.350 | 4.172 | 5.891 | 7.524 | 9.294 |
| 2006 | 0.304 | 1.108 | 2.338 | 4.099 | 5.845 | 7.457 | 9.225 |
| 2007 | 0.296 | 1.102 | 2.326 | 4.023 | 5.799 | 7.387 | 9.147 |


|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 0.288 | 1.096 | 2.315 | 3.944 | 5.752 | 7.313 | 9.061 |
| 2009 | 0.281 | 1.090 | 2.305 | 3.865 | 5.706 | 7.237 | 8.967 |
| 2010 | 0.273 | 1.084 | 2.298 | 3.790 | 5.660 | 7.159 | 8.865 |
| 2011 | 0.265 | 1.078 | 2.294 | 3.726 | 5.613 | 7.079 | 8.755 |
| 2012 | 0.258 | 1.072 | 2.295 | 3.678 | 5.567 | 6.997 | 8.640 |
| 2013 | 0.250 | 1.065 | 2.302 | 3.649 | 5.521 | 6.915 | 8.519 |
| 2014 | 0.242 | 1.059 | 2.317 | 3.643 | 5.475 | 6.834 | 8.395 |
| 2015 | 0.235 | 1.053 | 2.340 | 3.657 | 5.429 | 6.753 | 8.267 |
| 2016 | 0.227 | 1.047 | 2.373 | 3.690 | 5.383 | 6.673 | 8.137 |
| 2017 | 0.219 | 1.041 | 2.414 | 3.736 | 5.338 | 6.595 | 8.004 |
| 2018 | 0.212 | 1.035 | 2.462 | 3.791 | 5.292 | 6.518 | 7.870 |
| 2019 | 0.204 | 1.028 | 2.516 | 3.850 | 5.247 | 6.442 | 7.735 |
| 2020 | 0.196 | 1.022 | 2.574 | 3.912 | 5.202 | 6.367 | 7.599 |
| 2021 | 0.196 | 1.022 | 2.633 | 3.974 | 5.156 | 6.292 | 7.463 |

Table 3.10. Cod.27.6a. Natural mortality.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.496 | 0.378 | 0.298 | 0.256 | 0.231 | 0.218 | 0.207 |
| 1982 | 0.498 | 0.379 | 0.298 | 0.256 | 0.231 | 0.218 | 0.207 |
| 1983 | 0.501 | 0.379 | 0.298 | 0.256 | 0.232 | 0.218 | 0.207 |
| 1984 | 0.503 | 0.380 | 0.299 | 0.257 | 0.232 | 0.219 | 0.207 |
| 1985 | 0.505 | 0.380 | 0.299 | 0.257 | 0.232 | 0.219 | 0.208 |
| 1986 | 0.508 | 0.381 | 0.300 | 0.257 | 0.233 | 0.219 | 0.208 |
| 1987 | 0.510 | 0.382 | 0.300 | 0.258 | 0.233 | 0.219 | 0.208 |
| 1988 | 0.513 | 0.382 | 0.301 | 0.258 | 0.234 | 0.219 | 0.208 |
| 1989 | 0.515 | 0.383 | 0.301 | 0.258 | 0.234 | 0.220 | 0.208 |
| 1990 | 0.518 | 0.383 | 0.302 | 0.259 | 0.235 | 0.220 | 0.208 |
| 1991 | 0.521 | 0.384 | 0.302 | 0.259 | 0.235 | 0.220 | 0.208 |
| 1992 | 0.524 | 0.384 | 0.303 | 0.259 | 0.235 | 0.220 | 0.209 |
| 1993 | 0.527 | 0.385 | 0.304 | 0.259 | 0.236 | 0.221 | 0.209 |
| 1994 | 0.529 | 0.386 | 0.305 | 0.259 | 0.236 | 0.221 | 0.209 |
| 1995 | 0.532 | 0.386 | 0.306 | 0.259 | 0.237 | 0.221 | 0.209 |
| 1996 | 0.536 | 0.387 | 0.307 | 0.259 | 0.237 | 0.221 | 0.209 |
| 1997 | 0.539 | 0.387 | 0.308 | 0.260 | 0.238 | 0.222 | 0.210 |
| 1998 | 0.542 | 0.388 | 0.309 | 0.260 | 0.238 | 0.222 | 0.210 |
| 1999 | 0.545 | 0.388 | 0.311 | 0.261 | 0.239 | 0.223 | 0.210 |
| 2000 | 0.549 | 0.389 | 0.312 | 0.262 | 0.239 | 0.223 | 0.210 |
| 2001 | 0.552 | 0.390 | 0.313 | 0.263 | 0.240 | 0.223 | 0.211 |
| 2002 | 0.556 | 0.390 | 0.314 | 0.264 | 0.240 | 0.224 | 0.211 |
| 2003 | 0.560 | 0.391 | 0.315 | 0.265 | 0.241 | 0.224 | 0.211 |
| 2004 | 0.564 | 0.392 | 0.315 | 0.266 | 0.241 | 0.225 | 0.212 |
| 2005 | 0.568 | 0.392 | 0.316 | 0.267 | 0.242 | 0.225 | 0.212 |
| 2006 | 0.572 | 0.393 | 0.316 | 0.269 | 0.243 | 0.226 | 0.212 |
| 2007 | 0.576 | 0.393 | 0.317 | 0.270 | 0.243 | 0.227 | 0.213 |


|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 0.580 | 0.394 | 0.317 | 0.272 | 0.244 | 0.227 | 0.214 |
| 2009 | 0.585 | 0.395 | 0.318 | 0.273 | 0.244 | 0.228 | 0.214 |
| 2010 | 0.590 | 0.395 | 0.318 | 0.275 | 0.245 | 0.229 | 0.215 |
| 2011 | 0.595 | 0.396 | 0.318 | 0.276 | 0.245 | 0.229 | 0.216 |
| 2012 | 0.600 | 0.397 | 0.318 | 0.277 | 0.246 | 0.230 | 0.217 |
| 2013 | 0.605 | 0.397 | 0.318 | 0.278 | 0.247 | 0.231 | 0.217 |
| 2014 | 0.610 | 0.398 | 0.317 | 0.278 | 0.247 | 0.232 | 0.218 |
| 2015 | 0.616 | 0.399 | 0.316 | 0.278 | 0.248 | 0.233 | 0.219 |
| 2016 | 0.622 | 0.399 | 0.315 | 0.277 | 0.248 | 0.233 | 0.220 |
| 2017 | 0.628 | 0.400 | 0.313 | 0.276 | 0.249 | 0.234 | 0.221 |
| 2018 | 0.635 | 0.401 | 0.312 | 0.275 | 0.250 | 0.235 | 0.222 |
| 2019 | 0.642 | 0.401 | 0.310 | 0.274 | 0.250 | 0.236 | 0.224 |
| 2020 | 0.649 | 0.402 | 0.308 | 0.272 | 0.251 | 0.237 | 0.225 |
| 2021 | 0.649 | 0.402 | 0.306 | 0.271 | 0.252 | 0.237 | 0.226 |

Table 3.11. Cod.27.6a. Survey data made available to the WG. Data used in update assessment are highlighted in bold. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

ScoGFS- WIBTS- Q1:
Scottish west coast groundfish survey (ages 1-6 used)

| Effort (Hrs) | 1 | 2 | 3 | 4 | 5 | 6 | 7 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 | 1.5 | 23.7 | 8.6 | 13.6 | 3.9 | 2.5 | 1.2 | 1985 |
| 10 | 1.5 | 6.9 | 26.8 | 5.6 | 7.3 | 2.5 | 1.9 | 1986 |
| 10 | 57.4 | 16.2 | 15.3 | 22.8 | 3.0 | 2.8 | 0.0 | 1987 |
| 10 | 0.0 | 64.9 | 14.2 | 3.4 | 2.1 | 0.7 | 0.2 | 1988 |
| 10 | 4.5 | 7.2 | 45.1 | 8.6 | 1.9 | 0.5 | 0.8 | 1989 |
| 10 | 2.0 | 24.6 | 4.1 | 14.7 | 4.2 | 1.6 | 0.8 | 1990 |
| 10 | 4.8 | 5.4 | 17.4 | 5.2 | 13.4 | 2.8 | 0.5 | 1991 |
| 10 | 7.3 | 11.5 | 5.4 | 7.6 | 3.4 | 2.3 | 0.5 | 1992 |
| 10 | 1.7 | 38.2 | 12.7 | 1.7 | 1.4 | 1.1 | 0.0 | 1993 |
| 10 | 13.6 | 14.7 | 25.1 | 5.8 | 1.0 | 0.0 | 0.0 | 1994 |
| 10 | 6.4 | 23.8 | 14.0 | 16.5 | 1.2 | 1.9 | 0.7 | 1995 |
| 10 | 2.8 | 20.9 | 24.1 | 4.1 | 2.8 | 1.3 | 0.0 | 1996 |
| 10 | 11.1 | 7.7 | 11.6 | 7.9 | 4.2 | 4.7 | 1.0 | 1997 |
| 10 | 2.8 | 30.9 | 5.3 | 8.7 | 3.7 | 0.6 | 2.0 | 1998 |
| 10 | 1.5 | 8.2 | 8.2 | 1.4 | 3.2 | 0.5 | 0.5 | 1999 |
| 10 | 13.3 | 5.4 | 6.9 | 1.3 | 0.0 | 0.4 | 0.0 | 2000 |
| 10 | 2.7 | 18.4 | 5.7 | 13.2 | 19.5 | 1.1 | 1.6 | 2001 |
| 10 | 5.3 | 4.3 | 10.6 | 2.6 | 0.5 | 3.0 | 0.0 | 2002 |
| 10 | 2.7 | 16.7 | 2.0 | 4.7 | 1.8 | 0.7 | 0.4 | 2003 |
| 10 | 5.7 | 3.0 | 5.6 | 2.3 | 1.7 | 0.0 | 0.0 | 2004 |
| 10 | 1.3 | 1.5 | 1.2 | 0 | 0 | 0.4 | 0 | 2005 |
| 10 | 2.2 | 1.9 | 1.1 | 0.3 | 0 | 0 | 0.3 | 2006 |
| 10 | 2.1 | 18.8 | 3.4 | 1.2 | 0 | 0.6 | 0 | 2007 |
| 10 | 0.8 | 2.1 | 44.2 | 6.3 | 0.8 | 0 | 0 | 2008 |
| 10 | 1.8 | 2.6 | 2.3 | 0.4 | 0 | 0 | 0 | 2009 |
| 10 | 4.6 | 16.2 | 3.7 | 1.0 | 0.7 | 0 | 0 | 2010 |

Table 3.11. Continued. Cod.27.6a. Survey data made available to the WG. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

UK-SCOWCGFS-Q1 (index) (ages 1-6 used)

| Effort (Hrs) | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 | 0.52 | 32.95 | 21.07 | 0.93 | 0.98 | 0.74 | 0.00 | 2011 |
| 10 | 13.99 | 27.30 | 22.72 | 4.58 | 3.50 | 2.20 | 4.20 | 2012 |
| 10 | 20.03 | 40.26 | 26.38 | 36.95 | 7.76 | 0.30 | 0.00 | 2013 |
| 10 | 11.40 | 41.73 | 13.44 | 5.12 | 4.31 | 0.75 | 0.00 | 2014 |
| 10 | 8.16 | 36.40 | 70.70 | 37.74 | 23.25 | 13.00 | 2.47 | 2015 |
| 10 | 4.73 | 56.07 | 65.41 | 44.56 | 5.67 | 2.36 | 2.29 | 2016 |
| 10 | 2.92 | 33.49 | 50.58 | 49.58 | 156.64 | 10.71 | 24.89 | 2017 |
| 10 | 1.728 | 20.375 | 7.199 | 19.765 | 9.98 | 2.261 | 1.092 | 2018 |
| 10 | 9.924 | 4.173 | 6.888 | 2.031 | 3.181 | 0.318 | 0.318 | 2019 |
| 10 | 14.433 | 28.978 | 11.516 | 9.782 | 1.176 | 0.646 | 0.0 | 2020 |
| 10 | 1.175 | 12.137 | 22.988 | 2.946 | 2.519 | 1.236 | 0.0 | 2021 |

UK-SCOWCGFS-Q1 (variance)

| Effort (Hrs) | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 | 0.09 | 78.37 | 24.06 | 0.22 | 0.49 | 0.30 | 0.00 | 2011 |
| 10 | 44.18 | 120.08 | 33.73 | 2.31 | 8.34 | 4.83 | 13.02 | 2012 |
| 10 | 118.35 | 151.04 | 136.89 | 240.05 | 6.47 | 0.09 | 0.00 | 2013 |
| 10 | 20.17 | 383.27 | 12.23 | 3.04 | 5.47 | 0.28 | 0.00 | 2014 |
| 10 | 14.35 | 112.82 | 1264.73 | 602.27 | 289.82 | 98.91 | 5.48 | 2015 |
| 10 | 1.81 | 214.42 | 607.48 | 319.21 | 5.02 | 1.60 | 1.85 | 2016 |
| 10 | 1.43 | 155.67 | 498.57 | 1061.90 | 20475.95 | 84.79 | 287.62 | 2017 |
| 10 | 1 | 24.03 | 2.21 | 20.09 | 7.46 | 0.5 | 0.25 | 2018 |
| 10 | 6.79 | 2.03 | 6.12 | 0.6 | 1.98 | 0.1 | 0 | 2019 |
| 10 | 121.47 | 65.29 | 14.48 | 24.01 | 0.46 | 0.22 | 0 | 2020 |
| 10 | 1.03 | 10.19 | 31.36 | 1.35 | 0.92 | 0.37 | 0.13 | 2021 |

Table 3.11. Continued. Cod.27.6a. Survey data made available to the WG. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

| IreGFS | Irish groundfish survey |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 2002 |  |  |  |  |
| Effort (Hrs) | 0 | 1 | 2 | 3 |  |
| 1849 | 0.0 | 312.0 | 49.0 | 13.0 | 1993 |
| 1610 | 20.0 | 999.0 | 56.0 | 13.0 | 1994 |
| 1826 | 78.0 | 169.0 | 142.0 | 69.0 | 1995 |
| 1765 | 0.0 | 214.0 | 89.0 | 18.0 | 1996 |
| 1581 | 6.0 | 565.0 | 31.0 | 10.0 | 1997 |
| 1639 | 0.0 | 83.0 | 53.0 | 6.0 | 1998 |
| 1564 | 0.0 | 24.0 | 14.0 | 3.0 | 1999 |
| 1556 | 0.0 | 124.0 | 4.0 | 1.0 | 2000 |
| 755 | 3.0 | 82.0 | 28.0 | 2.0 | 2001 |
| 798 | 0.0 | 50.6 | 2.2 | 1.2 | 2002 |


| ScoGFS-WIBT |  | Quar | Scott | roun | sur | (ages | use |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 2010 |  |  |  |  |  |  |  |  |  |
| Effort (Hrs) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |  |
| 10 | 0 | 0.7 | 14.4 | 5 | 3 | 1.1 | 0.5 | 0 | 0 | 1996 |
| 10 | 1 | 10.9 | 2.4 | 1.4 | 1.4 | 1.4 | 0.2 | 0 | 0 | 1997 |
| 10 | + | 14.8 | 9.7 | 1.1 | 0 | 0 | 0 | 0 | 0 | 1998 |
| 10 | 2 | 4 | 6 | 9.2 | 0.5 | 0 | 0 | 0 | 0 | 1999 |
| 10 | 0 | 15.8 | 2.6 | 0.4 | 0.4 | 0 | 0 | 0 | 0 | 2000 |
| 10 | 1 | 1.7 | 7.3 | 1.7 | 0.3 | 0 | 0 | 0 | 0 | 2001 |
| 10 | 1 | 10.4 | 2.8 | 6.8 | 0.6 | 0 | 0 | 0 | 0 | 2002 |
| 10 | 1 | 1.5 | 11.3 | 2.9 | 0.6 | 0 | 0 | 0 | 0 | 2003 |
| 10 | 0 | 5.1 | 3.8 | 1.4 | 0 | 0.7 | 0 | 0 | 0 | 2004 |
| 10 | + | 2.1 | 3 | 0 | 0.6 | 0.3 | 0 | 0 | 0 | 2005 |
| 10 | 0 | 16.9 | 5.9 | 1.4 | 0.7 | 0 | 0 | 0 | 0 | 2006 |
| 10 | 0 | 12 | 20 | 1.3 | 0.5 | 0 | 0.3 | 0 | 0 | 2007 |
| 10 | 2 | 7.7 | 5 | 7 | 1 | 0 | 0 | 0 | 0 | 2008 |
| 10 | 2 | 14.2 | 3.8 | 1.2 | 1.2 | 0.3 | 0 | 0 | 0 | 2009 |
| 10 | na | na | na | na | na | na | na | na | na | 2010 |

Table 3.11. Cont. Cod.27.6a. Survey data made available to the WG. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

UK-SCOWCGFS-Q4 (index) (ages 1-6 used)

| Effort (Hrs) | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 10 | 0.60 | 9.71 | 31.54 | 10.88 | 0.93 | 1.70 | 2.38 | 0.00 | 0.00 | 2011 |
| 10 | 0.75 | 19.78 | 7.12 | 15.43 | 13.60 | 1.02 | 0.68 | 0.34 | 0.00 | 2012 |
| Survey not completed due to mechanical issues |  | 1.67 | 23.65 | 28.06 | 15.63 | 5.57 | 6.63 | 1.37 | 0.00 | 0.00 |
| 10 | 3.64 | 28.17 | 52.53 | 34.22 | 10.58 | 4.24 | 5.27 | 1.18 | 0.59 | 2014 |
| 10 | 0.374 | 6.162 | 34.941 | 45.443 | 118.92 | 14.893 | 5.773 | 3.176 | 0 | 2016 |
| 10 | 2.127 | 10.024 | 6.221 | 24.427 | 10.881 | 8.538 | 0.767 | 0.511 | 0 | 2017 |
| 10 | 0 | 4.569 | 15.945 | 4.809 | 39.902 | 29.022 | 10.887 | 0.829 | 0 | 2018 |
| 10 | 0.351 | 17.65 | 1.402 | 3.246 | 3.457 | 1.814 | 0.627 | 0.363 | 0 | 2019 |
| 10 | 0.801 | 15.988 | 24.873 | 3.472 | 4.936 | 1.35 | 0.783 | 0.392 | 0 | 2020 |
| 10 | 9.348 | 89.12 | 14.769 | 0.392 | 1.822 | 1.158 | 0.256 | 0 | 2021 |  |
| 10 | 0.863 |  |  |  |  |  |  |  |  |  |

UK-SCOWCGFS-Q4 (variance)

| Effort (Hrs) | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 10 | 0.21 | 31.08 | 38.07 | 5.78 | 0.19 | 1.56 | 4.79 | 0.00 | 0.00 | 2011 |
| 10 | 0.14 | 41.72 | 2.79 | 11.37 | 48.79 | 1.05 | 0.46 | 0.12 | 0.00 | 2012 |
| Survey not completed due to mechanical issues |  | 7.68 | 132.97 | 56.62 | 44.17 | 3.87 | 4.79 | 0.39 | 0.00 | 0.00 |
| 10 | 5.55 | 98.78 | 316.23 | 51.22 | 8.60 | 4.43 | 4.61 | 0.34 | 0.12 | 2014 |
| 10 | 0.14 | 7.394 | 419.36 | 716.38 | 7654.82 | 118.64 | 24.30 | 6.08 | 0 | 2016 |
| 10 | 3.215 | 11.252 | 3.816 | 76.154 | 14.262 | 8.928 | 0.207 | 0.063 | 0 | 2017 |
| 10 | 0 | 3.71 | 28.22 | 8.46 | 532.1 | 271.49 | 44.45 | 0.39 | 0 | 2018 |
| 10 | 0.03 | 88.63 | 0.43 | 1.86 | 2.6 | 0.67 | 0.39 | 0.13 | 0 | 2019 |
| 10 | 0.36 | 14.8 | 16.12 | 1.84 | 6.76 | 0.71 | 0.61 | 0.15 | 0 | 2020 |
| 10 | 0.25 | 9.38 | 4509.27 | 50.26 | 0.15 | 0.28 | 0.26 | 0.07 | 0 | 2021 |
| 10 |  |  |  |  |  |  |  |  |  |  |

Table 3.11. Continued. Cod.27.6a. Survey data made available to the WG. For the Scottish surveys, numbers are standardised to catch rate per ten hours. For the Irish surveys, effort is given as minutes towed and numbers are in units.

IRGFS-WIBTS-Q4 Irish West Coast groundfish. (ages 1-3 used)

| Effort (Hrs) | 0 | 1 | 2 | 3 | 4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1127 | 0 | 10 | 11 | 0 | 0 | 2003 |
| 1200 | 0 | 24 | 10 | 1 | 0 | 2004 |
| 960 | 63 | 13 | 7 | 0 | 2 | 2005 |
| 1510 | 0 | 95 | 12 | 0 | 0 | 2006 |
| 1173 | 0 | 161 | 12 | 0 | 1 | 2007 |
| 1135 | 0 | 23 | 24 | 4 | 0 | 2008 |
| 1378 | 1 | 75 | 4 | 5 | 0 | 2009 |
| 1291 | 0 | 70 | 31 | 4 | 3 | 2010 |
| 1287 | 1 | 26 | 26 | 4 | 0 | 2011 |
| 1230 | 0 | 74 | 7 | 3 | 0 | 2012 |
| 1295 | 0 | 92 | 11 | 0 | 0 | 2013 |
| 1200 | 0 | 113 | 20 | 2 | 0 | 2014 |
| 1213 | 0 | 15 | 11 | 3 | 0 | 2015 |
| 962 | 0 | 27 | 23 | 2 | 0 | 2016 |
| 1196 | 0 | 2 | 17 | 7 | 2 | 2017 |
| 966 | 1 | 21 | 3 | 0 | 1 | 2018 |
| 1291 | 0 | 36 | 1 | 0 | 0 | 2019 |
| 805 | 6 | 4 | 6 | 2 | 0 | 2020 |
| 1015 | 0 | 15 | 14 | 18 | 4 | 2021 |

Table 3.12. Cod.27.6a. SAM configuration file.
\# Where a matrix is specified rows corresponds to fleets and columns to ages.
\# Same number indicates same parameter used
\# Numbers (integers) starts from zero and must be consecutive
\#
\$minAge
\# The minimium age class in the assessment
1
\$maxAge
\# The maximum age class in the assessment
7
\$maxAgePlusGroup
\# Is last age group considered a plus group for each fleet (1 yes, or 0 no).
100000
\$keyLogFsta
\# Coupling of the fishing mortality states (nomally only first row is used).
$\begin{array}{lllllll}0 & 1 & 2 & 3 & 3 & 3 & 3\end{array}$
$\begin{array}{lllllll}-1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{cccccc}-1 & -1 & -1 & -1 & -1 & -1\end{array}-1$
$\begin{array}{cccccc}-1 & -1 & -1 & -1 & -1 & -1\end{array}-1$
-1 $-1 \begin{array}{lllll}1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{ccccccc}-1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$

## \$corFlag

\# Correlation of fishing mortality across ages ( 0 independent, 1 compound symmetry, $2 \operatorname{AR}(1), 3$ separable $\operatorname{AR}(1)$.
2
\$keyLogFpar
\# Coupling of the survey catchability parameters (normally first row is not used, as that is covered by fishing mortality).
-1 $-1 \begin{array}{lllll}1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{lllllll}0 & 1 & 2 & 3 & 4 & 5 & -1\end{array}$
$\begin{array}{llllll}6 & 7 & 7 & -1 & -1 & -1\end{array}$
8 9 1010 -1 - 1 - 1
$111213141515-1$
$161718192020-1$

## \$keyQpow

\# Density dependent catchability power parameters (if any).
-1 $-1 \begin{array}{lllll}1 & -1 & -1 & -1 & -1\end{array}$
-1 $-1 \begin{array}{lllll}1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{cccccc}-1 & -1 & -1 & -1 & -1 & -1\end{array}-1$
-1
-1 $-1 \begin{array}{lllll}1 & -1 & -1 & -1 & -1\end{array}$
$\begin{array}{lllllll}-1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
\$keyVarF
\# Coupling of process variance parameters for $\log (\mathrm{F})$-process (normally only first row is used)
$\begin{array}{lllllll}0 & 1 & 1 & 1 & 1 & 1\end{array}$

```
-1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1
-1 -1 -1 -1 -1 -1 -1
$keyVarLogN
# Coupling of process variance parameters for log(N)-process
0111111
```


## \$keyVarObs

```
\# Coupling of the variance parameters for the observations.
\(\begin{array}{lllllll}0 & 1 & 1 & 1 & 1 & 1 & 2\end{array}\)
\(\begin{array}{lllllll}3 & 3 & 3 & 3 & 3 & 3 & -1\end{array}\)
\(\begin{array}{llllll}4 & 4 & 4 & -1 & -1 & -1 \\ -1\end{array}\)
\(\begin{array}{llllll}5 & 5 & 5 & 5 & -1 & -1\end{array}-1\)
\(\begin{array}{lllllll}6 & 6 & 6 & 6 & 6 & 6 & -1\end{array}\)
\(777777-1\)
```


## \$obsCorStruct

\# Covariance structure for each fleet ("ID" independent, "AR" AR(1), or "US" for unstructured). I Possible values are: "ID" "AR" "US"
"ID" "ID" "ID" "ID" "ID" "ID"

## \$keyCorObs

\# Coupling of correlation parameters can only be specified if the $\operatorname{AR}(1)$ structure is chosen above.
\# NA's indicate where correlation parameters can be specified ( -1 where they cannot).
\#1-2 2-3 3-4 4-5 5-6 6-7
NA NA NA NA NA NA
NA NA NA NA NA -1
NA NA -1 -1 -1 -1
NA NA NA -1 -1 -1
NA NA NA NA NA -1
NA NA NA NA NA -1
\$stockRecruitmentModelCode
\# Stock recruitment code ( 0 for plain random walk, 1 for Ricker, 2 for Beverton-Holt, and 3 piece-wise constant). 0

## \$noScaledYears

\# Number of years where catch scaling is applied.
12

## \$keyScaledYears

\# A vector of the years where catch scaling is applied.
199519961997199819992000200120022003200420052006

## \$keyParScaledYA

\# A matrix specifying the couplings of scale parameters (nrow = no scaled years, ncols = no ages).
$\begin{array}{lllllll}0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$
$\begin{array}{lllllll}1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$

| 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 3 | 3 | 3 | 3 | 3 | 3 |
| 4 | 4 | 4 | 4 | 4 | 4 | 4 |
| 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| 6 | 6 | 6 | 6 | 6 | 6 | 6 |
| 7 | 7 | 7 | 7 | 7 | 7 | 7 |
| 8 | 8 | 8 | 8 | 8 | 8 | 8 |
| 9 | 9 | 9 | 9 | 9 | 9 | 9 |
| 10 | 10 | 10 | 10 | 10 | 10 | 10 |
| 11 | 11 | 11 | 11 | 11 | 11 | 11 |

## \$fbarRange

\# lowest and higest age included in Fbar
25
\$keyBiomassTreat
\# To be defined only if a biomass survey is used ( 0 SSB index, 1 catch index, 2 FSB index, 3 total catch, 4 total landings and 5 TSB index).
-1-1-1-1-1-1
\$obsLikelihoodFlag
\# Option for observational likelihood I Possible values are: "LN" "ALN"
"LN" "LN" "LN" "LN" "LN" "LN"
\$fixVarToWeight
\# If weight attribute is supplied for observations this option sets the treatment ( 0 relative weight, 1 fix variance to weight).
0
\$fracMixF
\# The fraction of $t(3)$ distribution used in $\log F$ increment distribution
0
\$fracMixN
\# The fraction of $\mathrm{t}(3)$ distribution used in $\log \mathrm{N}$ increment distribution
0

## \$fracMixObs

\# A vector with same length as number of fleets, where each element is the fraction of $t(3)$ distribution used in the distribution of that fleet

000000
\$constRecBreaks
\# Vector of break years between which recruitment is at constant level. The break year is included in the left interval. (This option is only used in combination with stock-recruitment code 3)

Table 3.13. Cod.27.6a. SAM estimated model parameters.

|  | par | sd(par) | $\exp (\mathrm{par})$ | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: |
| logFpar_0 | -9.82627 | 0.15868 | 0.00005 | 0.00004 | 0.00007 |
| logFpar_1 | -7.97290 | 0.15238 | 0.00034 | 0.00025 | 0.00047 |
| logFpar_2 | -7.09241 | 0.15248 | 0.00083 | 0.00061 | 0.00113 |
| logFpar_3 | -6.64846 | 0.15466 | 0.00130 | 0.00095 | 0.00177 |
| logFpar_4 | -6.17816 | 0.16649 | 0.00207 | 0.00149 | 0.00289 |
| logFpar_5 | -5.77349 | 0.17573 | 0.00311 | 0.00219 | 0.00442 |
| logFpar_6 | -11.07638 | 0.20538 | 0.00002 | 0.00001 | 0.00002 |
| logFpar_7 | -11.32750 | 0.16446 | 0.00001 | 0.00001 | 0.00002 |
| logFpar_8 | -8.21779 | 0.23178 | 0.00027 | 0.00017 | 0.00043 |
| logFpar_9 | -7.16734 | 0.23223 | 0.00077 | 0.00048 | 0.00123 |
| logFpar_10 | -6.87243 | 0.18456 | 0.00104 | 0.00072 | 0.00150 |
| logFpar_11 | -8.50034 | 0.22149 | 0.00020 | 0.00013 | 0.00032 |
| logFpar_12 | -6.33104 | 0.16249 | 0.00178 | 0.00129 | 0.00246 |
| logFpar_13 | -5.87522 | 0.16068 | 0.00281 | 0.00204 | 0.00387 |
| logFpar_14 | -5.46087 | 0.20059 | 0.00425 | 0.00285 | 0.00635 |
| logFpar_15 | -4.93287 | 0.20523 | 0.00721 | 0.00478 | 0.01086 |
| logFpar_16 | -6.96448 | 0.18865 | 0.00094 | 0.00065 | 0.00138 |
| logFpar_17 | -6.13471 | 0.14544 | 0.00217 | 0.00162 | 0.00290 |
| logFpar_18 | -5.52082 | 0.15834 | 0.00400 | 0.00292 | 0.00549 |
| logFpar_19 | -4.73993 | 0.20667 | 0.00874 | 0.00578 | 0.01321 |
| logFpar_20 | -4.14112 | 0.22248 | 0.01590 | 0.01019 | 0.02482 |
| logSdLogFsta_0 | -2.22062 | 0.73436 | 0.10854 | 0.02499 | 0.47148 |
| logSdLogFsta_1 | -2.37973 | 0.20440 | 0.09258 | 0.06151 | 0.13933 |
| $\operatorname{logSdLogN}$ _0 | -0.13522 | 0.12237 | 0.87353 | 0.68389 | 1.11575 |
| $\operatorname{logSdLogN\_ 1}$ | -2.53454 | 0.53318 | 0.07930 | 0.02730 | 0.23034 |
| logSdLogObs_0 | -0.54157 | 0.13520 | 0.58183 | 0.44398 | 0.76248 |
| logSdLogObs_1 | -1.51650 | 0.07853 | 0.21948 | 0.18758 | 0.25681 |


|  | par | sd(par) | $\exp (\mathrm{par})$ | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\operatorname{logSdLogObs\_ 2~}$ | -0.82221 | 0.12909 | 0.43946 | 0.33947 | 0.56891 |
| logSdLogObs_3 | -0.35243 | 0.06670 | 0.70298 | 0.61519 | 0.80329 |
| logSdLogObs_4 | $-0.16578$ | 0.10534 | 0.84723 | 0.68629 | 1.04592 |
| logSdLogObs_5 | $-0.24846$ | 0.11147 | 0.78000 | 0.62412 | 0.97481 |
| logSdLogObs_6 | 0.45734 | 0.09121 | 1.57987 | 1.31642 | 1.89604 |
| logSdLogObs_7 | 0.28239 | 0.10186 | 1.32629 | 1.08183 | 1.62598 |
| itrans_rho_0 | 0.90870 | 0.40524 | 2.48110 | 1.10320 | 5.57998 |
| logScale_0 | 0.02550 | 0.15205 | 1.02582 | 0.75685 | 1.39039 |
| logScale_1 | -0.16192 | 0.17277 | 0.85051 | 0.60203 | 1.20155 |
| logScale_2 | -0.10723 | 0.18484 | 0.89832 | 0.62070 | 1.30011 |
| logScale_3 | 0.13244 | 0.19076 | 1.14161 | 0.77951 | 1.67190 |
| logScale_4 | 0.18771 | 0.19398 | 1.20648 | 0.81853 | 1.77831 |
| logScale_5 | 0.36493 | 0.19735 | 1.44041 | 0.97066 | 2.13750 |
| logScale_6 | 0.70711 | 0.20145 | 2.02812 | 1.35556 | 3.03437 |
| logScale_7 | 0.60486 | 0.20010 | 1.83100 | 1.22711 | 2.73206 |
| logScale_8 | 1.12059 | 0.19596 | 3.06666 | 2.07231 | 4.53812 |
| logScale_9 | 1.36555 | 0.18682 | 3.91786 | 2.69638 | 5.69268 |
| logScale_10 | 1.20026 | 0.17694 | 3.32099 | 2.33120 | 4.73103 |
| logScale_11 | 0.68888 | 0.23380 | 1.99148 | 1.24767 | 3.17872 |

Table 3.14. Cod.27.6a. SAM estimates of population numbers-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 10776 | 19502 | 7001 | 1981 | 475 | 63 | 51 |
| 1982 | 24158 | 5364 | 7788 | 2593 | 728 | 184 | 44 |
| 1983 | 14133 | 11850 | 2224 | 2780 | 896 | 258 | 82 |
| 1984 | 24070 | 6612 | 4627 | 775 | 877 | 293 | 112 |
| 1985 | 10610 | 11370 | 2454 | 1467 | 227 | 254 | 121 |
| 1986 | 21633 | 4729 | 4108 | 749 | 379 | 63 | 102 |
| 1987 | 42929 | 10267 | 1766 | 1300 | 214 | 108 | 49 |
| 1988 | 7360 | 18877 | 3724 | 538 | 351 | 61 | 45 |
| 1989 | 21585 | 3157 | 6256 | 1139 | 155 | 100 | 31 |
| 1990 | 7974 | 9495 | 1079 | 1729 | 320 | 44 | 37 |
| 1991 | 11762 | 3384 | 3321 | 352 | 492 | 100 | 26 |
| 1992 | 21848 | 5012 | 1089 | 1004 | 108 | 140 | 38 |
| 1993 | 7886 | 9524 | 1693 | 310 | 289 | 33 | 55 |
| 1994 | 13771 | 3512 | 3297 | 490 | 91 | 85 | 27 |
| 1995 | 10200 | 6004 | 1244 | 1006 | 139 | 27 | 33 |
| 1996 | 4222 | 4471 | 1926 | 365 | 294 | 41 | 18 |
| 1997 | 17062 | 1829 | 1408 | 513 | 105 | 86 | 17 |
| 1998 | 5413 | 7413 | 548 | 369 | 144 | 30 | 30 |
| 1999 | 4314 | 2212 | 2164 | 141 | 105 | 43 | 18 |
| 2000 | 14621 | 1808 | 642 | 563 | 39 | 30 | 18 |
| 2001 | 4137 | 6079 | 558 | 173 | 160 | 11 | 14 |
| 2002 | 6975 | 1700 | 1860 | 150 | 45 | 43 | 7 |
| 2003 | 2303 | 2806 | 494 | 499 | 41 | 12 | 13 |
| 2004 | 3243 | 908 | 791 | 120 | 130 | 11 | 7 |
| 2005 | 2204 | 1274 | 262 | 196 | 30 | 31 | 4 |
| 2006 | 7212 | 887 | 409 | 62 | 45 | 7 | 9 |
| 2007 | 2419 | 3051 | 293 | 106 | 13 | 11 | 4 |


|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 1751 | 980 | 1066 | 75 | 25 | 3 | 3 |
| 2009 | 5474 | 734 | 345 | 264 | 16 | 6 | 2 |
| 2010 | 6453 | 2365 | 270 | 84 | 54 | 3 | 2 |
| 2011 | 2450 | 2828 | 904 | 69 | 18 | 12 | 1 |
| 2012 | 4188 | 1082 | 1162 | 241 | 15 | 4 | 3 |
| 2013 | 7473 | 1843 | 487 | 383 | 63 | 4 | 2 |
| 2014 | 6372 | 3357 | 811 | 174 | 115 | 18 | 2 |
| 2015 | 5925 | 2875 | 1548 | 292 | 57 | 39 | 7 |
| 2016 | 2226 | 2763 | 1338 | 593 | 97 | 20 | 16 |
| 2017 | 2012 | 994 | 1338 | 534 | 201 | 34 | 13 |
| 2018 | 924 | 922 | 442 | 514 | 176 | 67 | 16 |
| 2019 | 4183 | 396 | 411 | 163 | 153 | 55 | 26 |
| 2020 | 2876 | 1863 | 181 | 155 | 48 | 46 | 25 |
| 2021 | 1974 | 1264 | 878 | 70 | 46 | 14 | 22 |

Table 3.15. Cod.27.6a. SAM estimates for fishing mortality-at-age.

|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | 7 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1981 | 0.217 | 0.517 | 0.689 | 0.750 | 0.750 | 0.750 | 0.750 |
| 1982 | 0.226 | 0.531 | 0.722 | 0.811 | 0.811 | 0.811 | 0.811 |
| 1983 | 0.245 | 0.568 | 0.778 | 0.894 | 0.894 | 0.894 | 0.894 |
| 1984 | 0.264 | 0.607 | 0.844 | 0.988 | 0.988 | 0.988 | 0.988 |
| 1985 | 0.285 | 0.643 | 0.894 | 1.061 | 1.061 | 1.061 | 1.061 |
| 1986 | 0.279 | 0.329 | 0.303 | 0.302 | 0.362 | 0.369 | 0.859 |


|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 0.269 | 0.618 | 1.050 | 1.230 | 1.230 | 1.230 | 1.230 |
| 2009 | 0.260 | 0.599 | 1.076 | 1.301 | 1.301 | 1.301 | 1.301 |
| 2010 | 0.243 | 0.554 | 1.020 | 1.234 | 1.234 | 1.234 | 1.234 |
| 2011 | 0.229 | 0.513 | 0.968 | 1.198 | 1.198 | 1.198 | 1.198 |
| 2012 | 0.206 | 0.450 | 0.830 | 1.047 | 1.047 | 1.047 | 1.047 |
| 2013 | 0.193 | 0.415 | 0.749 | 0.951 | 0.951 | 0.951 | 0.951 |
| 2014 | 0.183 | 0.390 | 0.691 | 0.865 | 0.865 | 0.865 | 0.865 |
| 2015 | 0.173 | 0.367 | 0.641 | 0.816 | 0.816 | 0.816 | 0.816 |
| 2016 | 0.171 | 0.359 | 0.618 | 0.801 | 0.801 | 0.801 | 0.801 |
| 2017 | 0.179 | 0.380 | 0.653 | 0.861 | 0.861 | 0.861 | 0.861 |
| 2018 | 0.182 | 0.390 | 0.676 | 0.921 | 0.921 | 0.921 | 0.921 |
| 2019 | 0.179 | 0.384 | 0.675 | 0.965 | 0.965 | 0.965 | 0.965 |
| 2020 | 0.171 | 0.362 | 0.642 | 0.957 | 0.957 | 0.957 | 0.957 |

## Table 3.16. Cod.27.6a. SAM summary table. ('Catch' refers to model estimate).





Table 3.17. Cod.27.6a. Intermediate year assumptions based on the SAM assessment. Units are tonnes (SSB, landings, discards and catch) or thousands (recruitment).

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Fages 2-5 (2022) | 0.74 | $F_{\text {average (2019-2021) }}$ |
| SSB (2023) | 2923 | Short-term forecast; in tonnes. |
| Rage 1 (2022, 2023 and 2024) $^{2} 2226$ | Median recruitment resampled from the years 2016-2021; in thousands. |  |
| Total catch (2022) | 1873 | Short-term forecast; in tonnes. |
| Projected landings (2022) | 1333 | Short-term forecast assuming average landing pattern (2019-2021)^; in <br> tonnes. |
| Projected discards (2022) | 540 | Short-term forecast assuming average discard pattern (2019-2021)^; in <br> tonnes. |

${ }^{\wedge}$ Due to inadequate discard sampling coverage of the fishery in 2021, average landings and discards proportions from 2019-2020 are used for ages 1 and 2.

Table 3.18. Cod.27.6a. Catch scenarios based on the SAM assessment and assuming $F$ status quo in the intermediate year. Units are tonnes (SSB, landings, discards and catch) or thousands (recruitment).

| Basis Cat | Lan | Dis | Ftot | Flan | Fdis | SSB |  | SSB | TAC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY approach: F_MSY 809 | 594 | 215 | 0.3 | 0.23 | 0.069 | 3757 |  | 29\% | -37\% |
| Precautionary approach: 1359 Fpa | 985 | 374 | 0.57 | 0.44 | 0.13 | 3099 |  | 6.00\% | 6.30\% |
| FMSY upper 1210 | 881 | 329 | 0.49 | 0.38 | 0.112 | 3278 |  | 12.10\% | -5.40\% |
| FMSY lower 514 | 380 | 134 | 0.18 | 0.139 | 0.041 | 4113 |  | 41\% | -60\% |
| $\mathrm{F}=0 \quad 0$ | 0 | 0 | 0 | 0 | 0 | 4728 |  | 62\% | -100\% |
| Fpa 1359 | 985 | 374 | 0.57 | 0.44 | 0.13 | 3099 |  | 6.00\% | 6.30\% |
| F = Flim 1625 | 1169 | 456 | 0.73 | 0.56 | 0.167 | 2781 |  | -4.90\% | 27\% |
| Fsq 1642 | 1181 | 461 | 0.74 | 0.57 | 0.17 | 2760 |  | -5.60\% | 28\% |
| zero TAC advice - haddock 2562 Fmult | 1779 | 783 | 1.6 | 1.23 | 0.37 | 1650 |  | -44\% | 100\% |
| $\begin{aligned} & \hline \text { zero TAC advice - saith } 1642 \\ & \text { Fmult } \end{aligned}$ | 1181 | 461 | 0.74 | 0.57 | 0.17 | 2760 |  | -5.60\% | 28\% |
| $0.05 *$ Fsq 114 | 85 | 29 | 0.037 | 0.029 | 0.008 | 4590 |  | 57\% | -91\% |
| $0.25 *$ Fsq 527 | 390 | 137 | 0.185 | 0.143 | 0.042 | 4095 |  | 40\% | -59\% |
| $0.5^{*} \mathrm{Fsq}$ | 708 | 258 | 0.37 | 0.29 | 0.085 | 3568 |  | 22\% | -24\% |
| $0.75 *$ Fsq 1333 | 967 | 366 | 0.56 | 0.43 | 0.127 | 3130 |  | 7.10\% | 4.20\% |
| 2022F=Fsq then Fmsy HCR 133 | 99 | 34 | 0.043 | 0.033 | 0.01 | 4569 |  | 56\% | -90\% |
| 2022 F=Fsq then Fmsy HCR 80 lower | 60 | 20 | 0.026 | 0.02 | 0.006 | 4632 |  | 58\% | -94\% |
| 2022F=Fsq then Fmsy HCR 214 upper | 159 | 55 | 0.071 | 0.055 | 0.016 | 4472 |  | 53\% | -83\% |
| 2022F=Fsq then $0 \%$ SSB in- 1507 crease | 1088 | 419 | 0.66 | 0.51 | 0.15 | 2923 |  | 0.00\% | 17.80\% |
| 2022F=Fsq then $10 \%$ SSB in- 1263 crease | 918 | 345 | 0.52 | 0.4 | 0.119 | 3215 |  | 10.00\% | -1.25\% |
| 2022F=Fsq then 20\% SSB in-1019 crease | 745 | 274 | 0.4 | 0.3 | 0.09 | 3508 |  | 20\% | -20\% |
| $2022 \mathrm{~F}=$ Fsq then $30 \%$ SSB in- 773 crease | 569 | 204 | 0.28 | 0.22 | 0.065 | 3800 |  | 30\% | -40\% |
| $2022 \mathrm{~F}=\mathrm{Fsq}$ then $40 \%$ SSB in- 530 crease | 392 | 138 | 0.186 | 0.144 | 0.042 | 4092 |  | 40\% | -59\% |
| $2022 \mathrm{~F}=\mathrm{Fsq}$ then $50 \%$ SSB in- 286 crease | 212 | 74 | 0.096 | 0.074 | 0.022 | 4385 |  | 50\% | -78\% |
| $2022 \mathrm{~F}=\mathrm{Fsq}$ then prev. $\mathrm{TAC}^{*} 0.25$ | 237 | 83 | 0.108 | 0.083 | 0.025 | 4347 |  | 49\% | -75\% |
| $2022 \mathrm{~F}=\mathrm{Fsq}$ then prev.TAC 0.5 | 472 | 168 | 0.23 | 0.177 | 0.052 | 3961 |  | 36\% | -50\% |
| 2022F=Fsq then prev.TAC* 0.75 | 959 | 703 | 256 | 0.37 | 0.28 | 0.084 | 3577 | - 22\% | - $25 \%$ |
| $2022 \mathrm{~F}=\mathrm{Fsq}$ then prev.TAC ${ }^{*} 1$ | 1279 | 929 | 350 | 0.53 | 0.41 | 0.12 | 3196 | -9.30\% | 0.00\% |



Figure 3.1. Cod.27.6a. ICES official landings by country.


Figure 3.2. Cod.27.6a. ICES estimates of reported (red) and area misreported landings (blue) of cod caught in ICES Division 6.a.


Figure 3.3. Cod.27.6a. Amounts landed by métier (kg) in 2021 as submitted to InterCatch.


Figure 3.4. Cod.27.6a. Discard rates by weight by métier in 2021 as submitted to InterCatch.


Figure 3.5. Cod.27.6a. Discard rates after allocations within InterCatch.


Figure 3.6. Cod.27.6a. Landings (grey), imported (black) and raised (red, but so small so not visible) discards of all fleets after allocations within InterCatch.


Figure 3.7. Cod.27.6a. Catch numbers-at-age by sampled and unsampled landings and sampled and raised (unsampled) discards, after allocations within InterCatch.


Figure 3.8. Cod.27.6a. Catch numbers-at-age by fleet/catch category after allocations within InterCatch, 2018 (top left) to 2021 (bottom right).


Figure 3.9. Cod.27.6a. Landings and discards estimates by weight, as used by the WG.


Figure 3.10. Cod.27.6a. Discard proportion (of total catch) by weight.


Figure 3.11. Cod.27.6a. Discard proportion by number by age.


Figure 3.12. Cod.27.6a. Catch-at-age in numbers by year. Red: discards, blue: landings.


Figure 3.13. Cod.27.6a. Mean weights-at-age in landings and discards.


Figure 3.14. Cod.27.6a. Mean discard weights-at-age from Scottish sampling.


Figure 3.15. Cod.27.6a. Natural mortality-at-age based on stock weight-at-age and mortality-weight relationship (Lorenzen, 1996). (Age 1 bottom left, Age 7+ top right). Black: 2021, Grey: 2020.


Figure 3.16. Cod.27.6a. CPUE numbers for fish aged at 1+ per tow resulting from Scottish quarter one survey (UK-SCOWCGFS-Q1) in red and (UK-SCOWCGFS-Q4) in blue. Numbers are standardised to 30 minutes towing. Green polygons are areas closed to fishing.


Figure 3.17. Cod.27.6a. Log catch numbers-at-age (upper) and catch curves (lower) from commercial catch-atage data.


Figure 3.18. Cod.27.6a. Log catch (landings + discards) curve gradient plot using WG commercial catch-at-age data over different age ranges.


Figure 3.19. Cod.27.6a. Mean standardised catch-at-age proportions by number.


Figure 3.20. Cod.27.6a. Log mean standardised index values -by year- (left) and cohort (right) from Scottish quarter one ground fish survey (ScoGFS-WIBTS-Q1); ages 1-6. Survey finished in 2010.


Figure 3.21. Cod.27.6a. Log catch curves from Scottish quarter one ground fish survey (ScoGFS-WIBTS-Q1); ages 1-6. Survey finished in 2010.

## ScoGFS-WIBTS-Q1



Figure 3.22. Cod.27.6a. Within-survey correlations for the Scottish quarter one ground fish survey (ScoGFS-WIBTS-Q1), comparing index values at different ages for the same cohorts. The straight line in a linear regression. Survey finished in 2010.


Figure 3.23. Cod.27.6a. Log catch curve gradient plot using ScoGFS-WIBTS-Q1 index data. Solid line shows time-series of gradient of linear fit to curve over the age range $2 \mathbf{- 5}$, dashed line over the ages $2 \mathbf{- 4}$ and dotted line over the ages 3-5. Last cohort shown was at-age 5 in 2010, the last year of the ScoGFS-WIBTS-Q1 survey.


Figure 3.24. Cod in Division6a. Log mean standardised index values by year (left) and cohort (right) from ScoGFS-WIBTS-Q4.

log index

Figure 3.25. Cod.27.6a. Within survey correlations for ScoGFS-WIBTS-Q4 survey, comparing index values at different ages for the same cohorts. The solid line is a linear regression. Insufficient age 6 fish are caught to enable scatterplots to be constructed.


Figure 3.26. Cod.27.6a. Log mean standardised index values -by year (left) and cohort (right) from Irish quarter four ground fish survey (IRGFS-WIBTS-Q4); ages 1-3. Survey started in 2003.


Figure 3.27. Cod.27.6a. Log catch curves from Irish quarter four ground fish survey (IRGFS-WIBTS-Q4); ages 1-3. Survey started in 2003.


Figure 3.28. Cod.27.6a. Within-survey correlations for the Irish quarter four ground fish survey (IRGFS-WIBTS-Q4), comparing index values at different ages for the same cohorts. The straight line is a linear regression.


Figure 3.29. Cod.27.6a. Log mean standardised index values -by year (left) and cohort (right) - from Scottish quarter one ground fish survey UK-SCOWCGFS-Q1; ages 1-6.


Figure 3.30. Cod.27.6a. Log catch curves from new Scottish quarter one ground fish survey (UK-SCOWCGFSQ1); ages 1-7. Survey started in 2011.


Figure 3.31. Cod.27.6a. Within survey scatterplots from new Scottish quarter one ground fish survey (UK-SCOWCGFS-Q1), comparing index values at different ages for the same cohorts. The straight line in a linear regression.



Figure 3.32. Cod.27.6a. Log mean standardised index values by year (left) and cohort (right) from Scottish quarter four ground fish survey UK-SCOWCGFS-Q4); ages 1-6.


Figure 3.33. Cod.27.6a. Log catch curves from new Scottish quarter four ground fish survey (UK-SCOWCGFSQ4).

log index

Figure 3.34. Cod.27.6a. Within survey scatterplots from new Scottish quarter four ground fish survey (UK-SCOWCGFS-Q4), comparing index values at different ages for the same cohorts. The straight line in a linear regression.


Figure 3.35. Cod.27.6a. Comparison of survey indices by age. Irish Q4 survey (IRGFS.Q4) is compared to the current Scottish surveys (SCO.Q1=UK-SCOWCGFS-Q1 \& SCO.Q4=UK-SCOWCGFS-Q4). Values are mean standardised over the time period in common (2011-2021).


Figure 3.36. Cod.27.6a. Retrospective sensitivity analysis of SAM assessment results to lack of intermediate year survey data. Solid black line: WGCSE 2021 final assessment. Solid lines include intermediate year survey data. Dotted line exclude data.


Figure 3.37. Cod.27.6a. Retrospective sensitivity analysis of SAM assessment results to exclusion of age $\mathbf{1}$ and age $\mathbf{2}$ catch data. Black line: WGCSE 2022 final assessment. Solid lines include all catch data. Dotted line exclude age 1 and 2.


Figure 3.38. Cod.27.6a. SAM final run. Comparison of model estimated and observed log catch numbers-at-age.


Figure 3.39. Cod.27.6a. SAM final run. Comparison of model estimated and observed log index at age (ScoGFS-WIBTSQ1).


Figure 3.40. Cod.27.6a. SAM final run. Comparison of model estimated and observed log index at-age (IRGFS-WIBTS-Q4).


Figure 3.41. Cod.27.6a. SAM final run. Comparison of model estimated and observed log index at-age (ScoGFS-WIBTSQ4).


Figure 3.42. Cod.27.6a. SAM final run. Comparison of model estimated and observed log index at-age (UK-SCOWCGFSQ1).


Figure 3.43. Cod.27.6a. SAM final run. Comparison of model estimated and observed log index at-age (UK-SCOWCGFSQ4).


Figure 3.44. Cod.27.6a. SAM final run. One step ahead residuals for catch-at-age data and survey indices (upper panel) and process residuals (lower panel).


Figure 3.45. Cod.27.6a. SAM final run. Leave one out sensitivity analysis.


Figure 3.46. Cod.27.6a. Retrospective plots of final SAM run.


Figure 3.47. Cod.27.6a. Summary of the stock assessment. ICES estimated landings and discards shown in the upper left panel (catches from 1995-2006 (unshaded) are excluded from the assessment). Shaded areas (F and SSB) and error bars (recruitment) correspond to $95 \%$ confidence intervals.


Figure 3.48. Cod.27.6a. SAM final run. Stock-recruit relationship. Numbers indicate recruitment year.


Figure 3.49. Cod.27.6a. SAM forecast assuming Fsq in the intermediate year followed by zero catch (the proposed advice) in subsequent years.


Figure 3.50. Cod.27.6a. Percentage contribution to landings yield in 2021 and SSB in 2022 by recruitment year (not year class). Blue ('TRUE') indicates forecast assumption rather than an assessment model estimate.


Figure 3.51. Cod.27.6a. Historical assessment comparison plots. Final year recruitment in 2022 assessment is assumed (resampled from 2016-2021) rather than an assessment model estimate due to lack of intermediate year survey.

## 4 Cod (Gadus morhua) in Division 6.b (Rockall)

## Assessment in 2022

There is no assessment for this stock. Advice is provided on a triennial basis (last provided in 2020) according to the approach outlined below.

In 2020, the updated assessment and advice followed the agreed procedures for category 6.2.0 of ICES RGLIFE data-limited stock (DLS) advice rules (ICES, 2017a) as set out in the stock annex. For stocks without information on abundance or exploitation, ICES considers that a precautionary reduction of catches should be implemented unless there is ancillary information clearly indicating that the current level of exploitation is appropriate for the stock.

New survey information provided in 2020 indicated increased catch rates since 2018 and stable landings per unit of effort since 2016 implying that the current level of exploitation was appropriate for this stock. A precautionary buffer was therefore not applied for the advice provided in 2020.

## ICES advice applicable in 2021-2023

ICES advises that when the precautionary approach is applied, catches should be no more than 14 tonnes in each of the years 2021, 2022, and 2023.

## ICES advice applicable in 2018-2020

ICES advises that when the precautionary approach is applied, catches should be no more than 14 tonnes in each of the years 2018, 2019, and 2020. ICES cannot quantify the corresponding landings.

## ICES advice applicable in 2016-2017

ICES advises that when the precautionary approach is applied, landings should be no more than 17 tonnes in each of the years 2016 and 2017. ICES cannot quantify the corresponding total catches.

### 4.1 Data

Official landings data for cod in $6 . b$ are shown by nation in Table 4.1 and Figure 4.1. Total reported landings were 25.5 tonnes in 2021. Some updates were made to historical landings and the landings for 2018 and 2019 are now considered incomplete due to missing landings associated with national GDPR issues in some countries. In the past, official landings have shown very high interannual variation and it is not known whether these are a true reflection of removals. Official landings by subdivision are given in Table 4.2. Most landings are taken from Subdivision 6.b.2.

Landings data have been uploaded to InterCatch for 2021. In addition, some landings age compositions and discard data were also uploaded to InterCatch. In recent years, only limited discard data have been submitted to InterCatch for this stock. Discarded weight has been submitted for the Scottish demersal otter trawl fleet (OTB_DEF, the main fleet operating in the area) for the years since 2014; however, there is high interannual variability in the estimated discard rate for
this fleet (see below). Discard information has also been provided by Ireland for some years. Given the uncertainty surrounding these data, no estimates of total discards are provided.

|  | 2017 | 2018 | 2019 | $\mathbf{2 0 2 0}$ | 2021 |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Scottish OTB_DEF | $1.96 \%$ | $60.7 \%$ | $2.4 \%$ | $0.6 \%$ | $23.4 \%$ |
| Irish OTB_DEF | $0.63 \%$ | NA | $18.3 \%$ | NA | $17.8 \%$ |

## Commercial LPUE data

Irish and Scottish landings, effort and LPUE are presented in Figures 4.2 and 4.3 and Tables 4.3 and 4.4. Figure 4.2 shows a large decline in the Irish LPUE between 1995 and 2003 followed by relatively stable values at a level much lower than at the start of the time-series. In 2017, there was a large increase in effort for this fleet exceeding the previous time-series maximum. The recording of hours fished data in the log sheets is not mandatory for the Scottish fleet and consequently the data are incomplete. Scottish otter-trawl fleet data are therefore in units of $\mathrm{kg} / \mathrm{kWday}$. The Scottish time-series is much shorter and relatively noisier than the Irish timeseries. Whilst there were marked increases in LPUE in 2015 and 2016, given the magnitude of increase it seems unlikely to be completely attributable to an increase in stock size (an almost five-fold increase over two years) and LPUE has subsequently declined.

## Survey data

Two trawl surveys cover Division 6b: the Scottish Irish Anglerfish Megrim Industry-Science Survey (SIAMISS, G3745) which generally takes place in quarter 2, and the Scottish Rockall IBTS (Rock-WIBTS-Q3, G4436) which takes place in quarter 3. Neither surveys have particularly high catch rates of cod. Table 4.5 contains catch rates by age from the Rock-WIBTS-Q3 survey. The survey typically catches on a very truncated age range of fish with a complete lack of 0-group and very few fish above age 4 and therefore may not be a particularly good index for cod in Division 6b. Biomass estimates from the SIAMISS survey (Figure 4.4) are very uncertain, but the mean value shows an increase to 2018 and decrease in 2019. These estimates have not been updated since then.

### 4.2 References

ICES. 2017a. Advice basis. In Report of the ICES Advisory Committee, 2017. ICES Advice 2017, Book 1, Section 1.2.

## Table 4.1. Cod in Division 6.b (Rockall). Official catch statistics.

| Country | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 18 | - | 1 | - | 31 | 5 | - | - | - | 1 | - | - | - | - | - | - | - |
| France | 9 | 17 | 5 | 7 | 2 | - | - | - | - | - | - | - | - | - | - | - | + |
| Germany | - | 3 | - | - | 3 | - | - | 126 | 2 | - | - | - | 10 | 22 | 3 | 11 | 1 |
| Ireland | - | - | - | - | - | - | 400 | 236 | 235 | 472 | 280 | 477 | 436 | 153 | 227 | 148 | 119 |
| Norway | 373 | 202 | 95 | 130 | 195 | 148 | 119 | 312 | 199 | 199 | 120 | 92 | 91 | 55 | 52 | 85 | 152 |
| Portugal | - | - | - | - | - | - | - | - | - | - | - | - | - | 5 | - | - | - |
| Russia | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 7 |
| Spain | 241 | 1200 | 1219 | 808 | 1345 | - | 64 | 70 | - | - | - | 2 | 5 | 1 | 6 | 4 | 3 |
| UK (E\&W and NI) | 161 | 114 | 93 | 69 | 56 | 131 | 8 | 23 | 26 | 103 | 25 | 90 | 23 | 20 | 32 | 22 | 4 |
| UK (Scotland) | 221 | 437 | 187 | 284 | 254 | 265 | 758 | 829 | 714 | 322 | 236 | 370 | 210 | 706 | 341 | 389 | 286 |
| UK | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Total | 1023 | 1973 | 1600 | 1298 | 1886 | 549 | 1349 | 1596 | 1176 | 1097 | 661 | 1031 | 775 | 962 | 661 | 659 | 572 |

Table 4.1. Continued. Cod in Division 6.b (Rockall). Official catch statistics.

| Country | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020* | 2021* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | - | - | - | - | - | - | 3 | 5 | - | - | - | - | - | - | - | - | - | - |  |
| France | - | + | + | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | + |  |
| Germany | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| Ireland | 40 | 18 | 11 | 7 | 12 | 23 | 24 | 41 | 20 | 6 | 12 | 1 | 2 | 6 | 5 | 15 | 17 | ** | ** | 14 | 4 |
| Norway | 89 | 28 | 25 | 23 | 7 | 7 | 12 | 12 | 25 | 27 | 49 | 11 | 3 | + | 18 | 11 | 3 | 1 | 5 | 1 |  |
| Portugal | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| Russia | 26 | - | - | - | - | - | - | - | 1 | - | - | - | - | - | - | - | - | - | - | - |  |
| Spain | 1 | - | 6 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | + | + |  |
| UK (E\&W and NI) | 2 | 2 | 3 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| UK (Scotland) | 176 | 67 | 57 | 45 | 43 | 29 | 26 | 41 | 48 | 23 | 37 | 11 | 9 | - | - | - | - | - | - | - |  |
| UK | - | - | - | - | - | - | - | - | - | - | - | - | - | 10 | 18 | 37 | 38 | 49 | 51 | 27 | 21 |
| Total | 334 | 115 | 102 | 75 | 62 | 58 | 62 | 94 | 97 | 61 | 98 | 23 | 14 | 15 | 41 | 62 | 58 | 50** | 56** | 42 | 25 |

* Preliminary.
** Incomplete/missing due to national GDPR clauses.
$+<0.5$ tonnes.

Table 4.2. Cod in 6.b. Official landings by subdivision.

| Year | Subdivision 6b.1 | Subdivision 6b.2 |
| :--- | :---: | :---: |
| 2012 | 0.13 | 23.04 |
| 2013 | 0.39 | 13.11 |
| 2014 | 0.02 | 16.25 |
| 2015 | 13.23 | 27.77 |
| 2016 | 3.16 | 59.12 |
| 2017 | 2.52 | 54.55 |
| $2018^{\wedge}$ | 1.452 | 53.8 |
| $2019^{\wedge}$ | 2.4 | 14.5 |
| $2020^{\wedge \wedge}$ | 0.6 | 4.1 |
| $2021^{\wedge \wedge}$ |  | 62.5 |

${ }^{\wedge}$ Incomplete due to national GDPR clauses.
${ }^{\wedge \wedge}$ Incomplete as UK data only reported at division level.

Table 4.3. Cod in 6.b. Landings, effort and LPUE data from the Irish otter-trawl fleet.

| Year | Landings tonnes | Effort '000s Hrs | Lpue Kg/Hr |
| :---: | :---: | :---: | :---: |
| 1995 | 414.9 | 9.1 | 45.39 |
| 1996 | 402 | 7.2 | 55.68 |
| 1997 | 130.5 | 7.2 | 18.2 |
| 1998 | 207.1 | 7.3 | 28.23 |
| 1999 | 137.8 | 8.79 | 15.88 |
| 2000 | 101.1 | 9.9 | 10.23 |
| 2001 | 33.3 | 7.2 | 4.6 |
| 2002 | 16.2 | 2.6 | 6.18 |
| 2003 | 9.9 | 4.5 | 2.18 |
| 2004 | 6.9 | 2.2 | 3.08 |
| 2005 | 8.8 | 3.3 | 2.68 |
| 2006 | 22.2 | 5.9 | 3.76 |
| 2007 | 24.2 | 6.6 | 3.68 |
| 2008 | 41.6 | 9.9 | 4.21 |
| 2009 | 21.7 | 4.4 | 4.97 |
| 2010 | 7.5 | 3.3 | 2.3 |
| 2011 | 10.2 | 2.5 | 4.01 |
| 2012 | 1 | 3.2 | 0.31 |
| 2013 | 1.8 | 3.8 | 0.46 |
| 2014 | 5.6 | 4.2 | 1.34 |
| 2015 | 5.1 | 4.7 | 1.07 |
| 2016 | 16.4 | 6.2 | 2.65 |
| 2017 | 17.3 | 14.9 | 1.16 |
| 2018 | 13.3 | 11.8 | 1.13 |
| 2019 | 9.5 | 17.2 | 0.55 |
| 2020 | 13.3 | 15.1 | 0.88 |

Table 4.4. Cod in 6.b. Landings, effort and LPUE data from the Scottish TR1 (OTB_DEF) fleet.

| Year | Inds(t) | eff(kwdays) | Lpue(kg/kwday) |
| :---: | :---: | :---: | :---: |
| 2003 | 64.09 | 2504466 | 0.0256 |
| 2004 | 39.76 | 1842103 | 0.0216 |
| 2005 | 42.98 | 1217357 | 0.0353 |
| 2006 | 28.25 | 1011354 | 0.0279 |
| 2007 | 25.98 | 1060551 | 0.0245 |
| 2008 | 40.29 | 1124197 | 0.0358 |
| 2009 | 47.76 | 1631239 | 0.0293 |
| 2010 | 22.65 | 1744452 | 0.0130 |
| 2011 | 36.54 | 1565753 | 0.0233 |
| 2012 | 10.78 | 901552 | 0.0120 |
| 2013 | 9.09 | 532767 | 0.0171 |
| 2014 | 9.70 | 668665 | 0.0145 |
| 2015 | 19.92 | 563098 | 0.0354 |
| 2016 | 34.01 | 514486 | 0.0661 |
| 2017 | 37.71 | 794571 | 0.0475 |
| 2018 | 49.25 | 794017 | 0.062 |
| 2019 | 50.65 | 1078714 | 0.046 |
| 2020 | 26.736 | 963518 | 0.027 |
| 2021 | 21.46 | 919033 | 0.023 |

Table 4.5. Cod in 6.b. Survey data made available to the WG: Scottish Q3 ground fish survey ((Rock-WIBTS-Q3)). Catch rates are given as number per 10 hours.

| Year | $\begin{aligned} & \text { Effort (10 } \\ & \text { Hours) } \end{aligned}$ | $\begin{aligned} & \text { Age } \\ & 0 \end{aligned}$ | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 10 | 0 | 0.493 | 0.493 | 0 | 0 | 0 | 0 | 0 | 0.403 | 0 |
| 2014 | 10 | 0 | 0.279 | 0.894 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 10 | 0 | 0 | 0.922 | 0.307 | 0 | 0 | 0 | 0 | 0 | 0.307 |
| 2016 | 10 | 0 | 0 | 0.269 | 0.538 | 0.538 | 0 | 0 | 0.269 | 0 | 0 |
| 2017 | 10 | 0 | 0 | 0 | 0 | 0.922 | 1.062 | 0 | 0 | 0 | 0 |
| 2018 | 10 | 0 | 0 | 0.307 | 0.614 | 0.307 | 0.307 | 0 | 0 | 0 | 0 |
| 2019 | 10 | 0 | 1.249 | 0.453 | 0.969 | 0.094 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 10 | 0 | 1.117 | 0.922 | 2.432 | 0.307 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 10 | 0 | 0 | 1.203 | 0.896 | 0.614 | 0 | 0 | 0 | 0 | 0 |



Figure 4.1. Cod in Division 6.b. Total of official landings by nation. In some cases, official landings are missing due to national GDPR clauses and in which case ICES estimates are presented here instead.


LPUE


Figure 4.2. Cod in Division 6.b. Landings, effort and LPUE (kg/hr) from the Irish Otter-trawl fleet. No update for 2021.


## Landings



Figure 4.3. Cod in Division 6.b. Landings, effort and LPUE (Kg/kWday) from the Scottish OTB_DEF fleet.


Figure 4.4. Cod in Division 6b. Estimated biomass from SIAMISS survey.

## 5 Cod in 7.a (Irish Sea)

Situated between Ireland and Great Britain the Irish Sea (7.a) is connected by to the Celtic Sea (7.g) at its southern extreme by the St George's Channel and in north is linked to sea region West of Scotland (6.a) by the Northern Channel. The average depth is 50 m but the area is contrasted between a deeper channel, in the west, and shallower bays in the east. The channel has a maximum depth exceeding 275 m whilst the eastern bays have depths less than 50 m . Distinct habitat patches result from a combination of bathymetry, topographical features and hydrography. The sea bed of the eastern Irish Sea is dominated by fine sediment plains with some small areas of areas of mud habitat, the fine sediments graduate to more coarse material in central areas. A large well-defined deep-water mud basin is located in the northwestern region in close to the Northern Irish and Irish coast.

Irish Sea fisheries are predominantly demersal trawling and seining with demersal trawling for Nephrops dominating effort with vessels using mesh in the range $70-99 \mathrm{~mm}$. Effort using fishing gear with $\geq 100 \mathrm{~mm}$ mesh sizes is currently at a low level compared to historic activity, a considerable decline in effort was observed between 2003 and 2007 and has continued since. The species composition of catches by vessels in using $\geq 100 \mathrm{~mm}$ mesh consists primarily of haddock, with lower quantities of hake. At present there is no commercial towed gear fishery for cod permitted. Beam trawls are operating within the Irish Sea with mesh sizes in the range $80-119 \mathrm{~mm}$, targeting sole, plaice, and rays. A seasonal pelagic and gillnet herring fishery operates in late summerearly autumn in the pre and post-spawning period. Dredge fisheries target king and queen scallops, with king scallops in coastal areas with the queen scallop fishery operating in the central area south of the Isle of Man, to a lesser extent queen scallops are also targeted using trawl nets, during the late summer when swimming activity is most pronounced.
There is a recreational fishery which catches cod and with declining commercial rates has become a more important aspect of the total catch. At the last benchmark in February 2022 (ICES, 2022a) the recreational fishery was included in the assessment for the first time.

## Type of assessment

The stock was benchmarked in February 2022 (ICES, 2022a) and a Stock Synthesis (SS3) fully analytical model is now being used in the cod assessment.

## ICES advice applicable to 2021 and 2022

ICES advised on the basis of precautionary approaches that there should be no directed fisheries, and bycatch and discards should be minimized in 2021 and 2022. Advice since 2020 was applied based on the on the 2 over 3 rule for category 3 .

## ICES advice applicable to 2023

ICES advised on the basis of the MSY and precautionary approach that there should be zero catches in 2023 as SSB will be below Blim in 2023 and 2024.

### 5.1 General

## Stock description and management units

The stock and the management unit are both ICES Division 7.a (Irish Sea).

## Management applicable to 2021

TACs and quotas set for 2021

|  | TAC | Landed |
| :--- | :---: | :---: |
| Belgium | 3 | 2.8 |
| France | 7 | 0 |
| Ireland | 104 | 41.8 |
| The Netherlands | 1 | 0 |
| United Kingdom | 91 | 88.7 |
| EU (total) | 115 | 44.6 |
| Total | 206 | 133.3 |

Management of this cod fishery is by TAC, days-at-sea limits and technical measures. Technical regulations in force in the Irish Sea, including those associated with the cod recovery plan since 2000, are described in Section 5.2.

Quota uptake in 2021 was considerably below the officially set TAC of 206 tonnes.
Table 1. Fishing opportunities (TAC) for 2022 for cod in 27.7.a.

| Species: | Cod Gadus morhua |  | Zone: | 7a (COD/07A.) |
| :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 | (1) | Precautionary TAC <br> Article 7(1) of this Regulation applies |  |
| France | 7 | (1) |  |  |
| Ireland | 104 | (1) |  |  |
| Netherlands | 1 | (1) |  |  |
| Union | 115 | (1) |  |  |
| United Kingdom | 91 | (1) |  |  |
| TAC | 206 | (1) |  |  |

(1) Exclusively for bycatches. No directed fisheries are permitted under this quota.
https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX\%3A32021R1239.

## Fishery in 2021

Landings in accordance with TAC were below the TAC, however, the TAC in 2021 was considerably higher than the ICES advised value of 74 tonnes.

Since 2009, Irish landings of cod reported from ICES rectangles immediately north of the Irish Sea/Celtic Sea boundary (ICES rectangles 33E2 and 33E3) have been reallocated into the Celtic Sea as they represent a combination of inaccurate area reporting and catches of cod considered by ICES to be part of the Celtic Sea stock (ICES, 2009). The amount of Irish landings transferred
from 7 a to $7 \mathrm{e}-\mathrm{k}$ by year is shown below. Total official landings for this stock in 2021 were 133 tonnes after this re-allocation and total catches in the area were 184 t .

Table 2. Transfers from ICES rectangles 33E2 and 33E3.

| Year | Tonnes |  |
| :---: | :---: | :---: |
|  | 2004 | 108 |
|  | 2005 | 54 |
|  | 2006 | 103 |
|  | 2007 | 527 |
|  | 2008 | 558 |
|  | 2009 | 193 |
|  | 2010 | 143 |
|  | 2011 | 147 |
|  | 2012 | 130 |
|  | 2013 | 75 |
|  | 2014 | 24 |
|  | 2015 | 39 |
|  | 2016 | 40 |
|  | 2017 | 19 |
|  | 2018 | 20 |
|  | 2019 | 37 |
|  | 2020 | 71 |
|  | 2021 | 52 |

The majority of landings in 2021 was taken by the Nephrops fleet, followed by the midwater demersal fleet. Landings and discards by métier and country can be seen in Table 8. Total uptake of cod TAC was $65 \%$.

A Fishery-Science Partnership Survey (FSP) was repeated in the western Irish Sea in spring 2021 in the western Irish Sea using semi-pelagic gear on commercial vessels. This survey attempts to address the lack of sampling opportunities created by the diminishing TAC for cod in the Irish Sea and the resulting significant reduction of a directed whitefish fleet targeting cod.

## InterCatch procedure

Since 2013 international landings and discards-at-age are uploaded into InterCatch. Discards are raised for unreported strata and métiers to estimate total discards-at-age.

## Landings

The input data on fishery landings and age compositions are split into four periods:

1. 1968-1990. Landings in this period, provided to ICES by stock coordinators from all countries, are assumed to be un-biased and are used directly as the input data to stock assessments.
2. 1991-1999. TAC reductions in this period caused substantial misreporting of cod landings into several major ports in one country, mainly species misreporting. Landings into these ports were estimated based on observations of cod landings by different fleet sectors during regular port visits. For other national landings, the WG figures provided to ICES stock coordinators were used.
3. 2000-2005. Cod recovery measures were considered to have caused significant problems with estimation of landings. The ICES WG landings data provided by stock coordinators for all countries are considered uncertain and estimated within an assessment model. Observations of misreported landings were available for 2000, 2001, 2002 and 2005. However, they have generally not been used to correct the reported landings but have been used to evaluate model estimates in those years.
4. Since 2006. The introduction of the UK buyers and sellers legislation is considered to have reduced the bias in the landings data but the level to which this has occurred is unknown. Consequently comparisons were made between the fit of the model to recorded landings under an assumption of bias and unbiased information.
5. 2020. The Covid-19 pandemic made the collection of observer data aboard vessels impossible for Q2-Q4, making the estimation of discard data and the establishment of age structure in catches impossible for most of the year. Age structure of the stock is available from Q1 observer data and the 3 surveys, FSP, and Q1 and Q4 groundfish surveys.
1. 2021. The continued COVID-19 situation resulted in reduced sampling; for the quarter 1 2021 the full final tow of the TR1 fleet was landed and sampled by observers ashore. There was very low sampling of cod in the Nephrops directed fleet, particularly in quarter 1 due to no observed trips. A raising procedure similar to the previous year was applied, in which the cod sampled in the Northern Ireland fishers self-sampling scheme were applied and raised to the full Nephrops catches. However, no cod were found in the provided self samples.

The annual numbers-at-age caught and the mean weights-at-age in landings (applied to the total catch) by age are given in Table 9 and Table 10; numbers of catch-at-age for 2020 are excluded due to limited discard and port sampling during the COVID-19 pandemic.

## Discards data

The WKIrish3 (ICES, $2017 \mathrm{a}, \mathrm{b}$ ) benchmark report gives details on historic raising to total national and international discards.

## Biological data

## Natural mortality

Natural mortality has been revised in WKNSCS (ICES, 2022a). M-at-age was calculated from tagging data following calculated following (Pollock, Hoenig et al., 1989, Hoenig, Barrowman et al., 1998). Natural mortality is kept constant throughout years.

## Maturity

Maturity ogive has been revised in WKIrish2 (ICES, 2016). Each year the smoother is applied to the full time-series of raw data and values are accordingly updated. Updated values after application of the smoother are in Table 12. Please refer to the stock annex for further information.

## Survey data used for advice

Please refer to the stock annex for a description of the surveys and survey data.

| Survey | Ages | Years |
| :--- | :---: | :---: |
| FSP SURVEY (B7897) | $2-6$ | $2004-2021$ (EXCLUDING 2014) |
| NIGFS-WIBTS-Q4 (G7655) | 0 | $1995-2021$ |
| NIGFS-WIBTS-Q1 (G7144) | $1-4$ | $1995-2021$ |

### 5.2 Historical stock development

The advice is based on the newly benchmarked assessment (WKNSCS, ICES 2022a).

## Deviations from Stock Annex

During the benchmark process kept and assumed dead (i.e. $35 \%$ of released fish) removals from the recreational fishery were included in the assessment. There was considerable discussion regarding the introduction of the recreational fishery, which was in the range of 30 tonnes for each of the available data year 2017-2020. The benchmark agreed and reviewed a) a model excluding the recreational data due to issues with uncertainty of data and insufficient information of a selectivity pattern and b) combine the recreational removals with the total commercial catches, using the same selectivity pattern. With a view on a possible increase of the recreational component over the next few years with decreasing commercial fishing pressure, the benchmark decided that it would be beneficial to go for approach $b$ to enable the incorporation of future recreational removals. The differences in SSB, Fbar and general perception of the stock were negligible.

In preparation for the working-group it became apparent that the values estimated for the recreational fleet ahead of the benchmark contained a data error and are now indeed estimated at approximately 120-150 tonnes for each of the year 2017-2021, similar levels to the commercial catches.

There are high uncertainties around the recreational removals and the values largely diverge from the benchmarked values.

The benchmark had agreed on a model excluding the recreational catches and passed the model as fit to benchmark and considered the inclusion of the recreational data only as a way to include them in times of increase in relation to commercial catches. In light of the data error the WGCSE working group therefore decided in this instance to exclude all recreational removals from the model.

## Final assessment

The final assessment has been run in stock Synthesis (SS3). Available data and catch-at-age, dis-cards-at-age and numbers-at-age in surveys can be seen in Figure 1-3, Figure 6 and Table 9-13, while summary of assessment results can be seen in Table 14.

The fit of the model catch-at-age data and to the indices is good, showing "all green" runs tests (Figure 11 and Figure 12) as well as the individual residuals. Further details on the use of the Runs tests and RMSE can be found in Carvalho et al., 2021. The retrospectives provide a good fit with Mohns Rho for SSB and $\mathrm{F}_{\mathrm{bar}}$ at 0.09 and -0.14 respectively (Figure 13).

The final results of the assessment can be seen in Figure 4-6.

## Final assessment: long-term trends

### 5.3 Short-term predictions

Short-term forecast was carried out in using the FLR forecast environment. Assumptions for the intermediate year can be seen in Table 3. Geometric mean for recruitment is from 2002-2019 (Final year-2), which encompasses the block where recruitment is supposedly reduced.

Table 3. Short-term forecast assumptions.

| Variable | Value | Notes |
| :--- | :---: | :--- |
| $F_{\text {ages 2-4 (2022) }}$ | 0.038 | $F_{\text {sq }}=F_{\text {average (2018-2021) }} *$ |
| SSB (2023) | 4842 | Short-term forecast fishing at $f_{\text {sq; in tonnes. }}$ |
| $R_{\text {age } 0}(2022$ and 2023) | 17989 | Geometric Mean (2002-2019); in thousands |
| Total catch (2022) | 165 | Fishing at $F_{\text {sq; in tonnes }}$ |
| Projected landings (2022) <br> ((2022)((20(2022(2020) | 159 | Assuming average landing patterns (2019-2021); in <br> tonnes |
| Projected discards (2022) | 6 | Assuming average discard patterns (2019-2021); in <br> tonnes |

* $F$ in 2020 was assumed to be unrepresentative due to the COVID-19 disruption and hence $F_{s q}$ was calculated as $F_{\text {average }}$ (2018-2021) excluding 2020.

Table 4 shows the catch scenarios, in particular the zero catch advice and the scaled MSY advices due to SSB being below MSY Btrigger and unable to reach Blim even under a no-catch scenario. The newly introduced Feco (ICES, 2022a) is also included in a scaled version. With the Sea Surface temperature Index being being high for the recent years, FECO is currently set at 0.19.

Table 4. Catch scenarios for 2023; all weights are in tonnes.

| Basis | Total catch (2023) | Projected landings (2023) | Pro- <br> jected <br> dis- <br> cards <br> (2023) | $\begin{aligned} & F_{\text {total }} \\ & (2023) \end{aligned}$ | $F_{\text {projected }}$ <br> landings <br> (2023) | $F_{\text {projected dis- }}$ cards (2023) | $\begin{aligned} & \text { SSB } \\ & \text { (2024) } \end{aligned}$ | \% SSB change | \% TAC change $\wedge$ | \% Advice change ${ }^{\wedge \wedge}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |  |  |
| MSY approach: $F=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 5410 | 11.7 | -100 | -100 |
| Other scenarios |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \mathrm{F}_{\text {MSY }} \times \text { SSB } \\ & (2023) / \text { MSY } \text { Brrig. }^{\text {ger }} \\ & \text { g. } \end{aligned}$ | 403 | 382 | 21 | 0.093 | 0.090 | 0.0036 | 4988 | 3.0 | 96 | 440 |
| $\begin{aligned} & \text { FMSY lower } \times \text { SSB } \\ & (2023) / \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |
| $F=\mathrm{F}_{\text {MSY }}$ lower | 704 | 667 | 37 | 0.168 | 0.162 | 0.0065 | 4677 | -3.4 | 240 | 850 |
| $F=F_{\text {MSY }}$ | 908 | 861 | 48 | 0.22 | 0.21 | 0.0086 | 4466 | -7.8 | 340 | 1130 |
| $F=F_{p a}$ | 1011 | 958 | 53 | 0.25 | 0.24 | 0.0097 | 4362 | -9.9 | 390 | 1270 |
| $\mathrm{F}=\mathrm{F}_{\text {MSY }}$ upper | 1093 | 1035 | 58 | 0.27 | 0.26 | 0.0106 | 4278 | -11.7 | 430 | 1380 |
| $F=F_{2022}$ | 170 | 161 | 9 | 0.038 | 0.037 | 0.00148 | 5232 | 8.0 | -17.5 | 130 |
| $\mathrm{F}=\mathrm{F}_{\text {lim }}$ | 1612 | 1526 | 86 | 0.43 | 0.41 | 0.0166 | 3754 | -22 | 680 | 2100 |
| $\mathrm{F}=\mathrm{F}_{\text {Eco }}$ | 788 | 747 | 41 | 0.19 | 0.183 | 0.0074 | 4589 | -5.2 | 280 | 960 |
| $\begin{aligned} & \mathrm{F}_{\mathrm{ECO}} \times \operatorname{SSB} \text { (2023) } \\ & / \mathrm{MSY} \mathrm{~B}_{\text {trigger }} \end{aligned}$ | 347 | 329 | 18 | 0.080 | 0.077 | 0.0031 | 5047 | 4.2 | 68 | 370 |
| $\begin{aligned} & \text { SSB (2024) = } \\ & \text { SSB (2023) } \end{aligned}$ | 544 | 516 | 28 | 0.128 | 0.123 | 0.0049 | 4842 | 0 | 160 | 640 |
| $\underset{* *}{\operatorname{SSB}(2024)}=\text { Blim }^{2}$ |  |  |  |  |  |  |  |  |  |  |

* SSB 2024 relative to SSB 2023.
** The Blim option was left blank because Blim cannot be achieved in 2024, even with zero catches.
${ }^{\wedge}$ Total TAC in 2023 relative to the TAC in 2022 ( 206 tonnes).
$\wedge \wedge$ Total Advice in 2023 relative to advice in 2022 ( 74 tonnes).


### 5.4 Biological reference points

New reference points were defined at WKNSCS (ICES, 2022a). The newly introduced Feco (ICES, 2022a) has been agreed and reviewed at the benchmark for a stock for the first time. FECO is an opportunity to use environmental data in forecast scenarios (ICES, 2022a). In case of cod in 7 .a a sea surface temperature (SST) was found to be a reasonable indicator for productivity. The Feco reference point uses the inverted SST (with a 3-year lag to account for the time from larvae stage to contribution to SSB) rescaled between zero and one which informs the status of the indicator
(Is) in the advice year compared with previous years. The status of the indicator determines the placement of the FECO reference point within FMSY ranges (ICES, 2019; 2020); for 2023 FECO is at 0.19 , estimated as FMSY lower + ((FMSY upper-FMSY lower)*Is).

Table 5. Biological reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $B_{\text {triger }}$ | 11538 | $\mathrm{B}_{\mathrm{pa}}$ | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.222 | Median point estimates of ( $\mathrm{F}_{\text {MSY }}$ ) EqSim with combined SR | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\text {MSY lower }}$ | 0.168 | Median lower estimates of ( $\mathrm{F}_{\text {MSY }}$ ) EqSim with combined SR | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\text {MSY upper }}$ | 0.273 | Median upper point estimates of ( $\mathrm{F}_{\text {MSY }}$ ) EqSim with combined SR | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\text {ECO }}$ | 0.19 | $\begin{aligned} & \text { Ecosystem Indicator }\left(I_{s}\right) ; \mathrm{F}_{\mathrm{ECO}}=\mathrm{F}_{\mathrm{MSY} \text { lower }}+\left(\left(\mathrm{F}_{\mathrm{MSY} \text { upper }}-\right.\right. \\ & \left.\left.\mathrm{F}_{\text {MSY lower }}\right)^{*} I_{s}\right) \end{aligned}$ | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
| Precautionary approach | $\mathrm{Bl}_{\text {lim }}$ | 8303 | Lowest SSB with above-average recruitment | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 11538 | Blim combined with the assessment error | ICES, 2022a |
|  | $\mathrm{F}_{\text {lim }}$ | 0.43 | F with 50\% probability of SSB less than $\mathrm{Bl}_{\text {lim }}$ | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.25 | $\mathrm{F}_{\text {P05 }}$; the F that leads to $\mathrm{SSB}>\mathrm{B}_{\text {lim }}$ with $95 \%$ probability | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |

### 5.5 Management plans

The Irish Sea cod management plan, as described in Council Regulation (EC) 1342/2008 was evaluated independently by ICES in 2009 using the approach adopted in AGCREMP 2008 and found to be not consistent with the ICES Precautionary Approach (WGCSE 2009).

### 5.6 Uncertainties and bias in assessment

## Surveys

The Irish Sea has relatively good survey coverage. The quarter 1 groundfish survey and the FSP survey have got good consistent cover of the age contributions. The Q 4 groundfish survey only attributes to the recruitment at age 0 .

## Stock structure and migrations

Stock structure and migrations have been in full discussed in the WKIrish2 report (ICES, 2016), however, there are still uncertainties and discussions.

A tagging study of Irish Sea cod and Celtic Sea cod was conducted from 2016-2019 in part to address these issues. Up to January 2019, 4238 cod were caught and tagged aboard chartered commercial fishing vessel using semi-pelagic fishing gear, FSP survey, shore angling competitions and others. Up to January 2019, 138 tagged cod were returned. The project relies on collaboration with the fishing industry to provide the data to develop a better understanding of the current behaviour, biology and stock status of Irish Sea cod. Most recent results suggest a stronger migratory behaviour of Irish Sea cod into the Celtic Sea, indicating that up to $18 \%$ of mature fish might leave the Irish Sea (ICES, 2021). This will have considerable impacts on the future management and assessment of the stock, but additional research is necessary. Currently a further project using data storage tags and trace element analysis is being conducted to understand stock structure and migratory behaviour as well as mixing.

### 5.7 Management considerations

A number of emergency and cod recovery plan measures have been introduced since 2000 to conserve Irish Sea cod. These include a spawning closure since 2000 and effort control since 2003. There have also been several vessel decommissioning schemes. As it has not been possible to provide analytical catch forecasts in recent years, the TAC has been reduced by 15-20\% annually since 2006 and by $25 \%$ since 2009. An MSY approach was used to set TAC in 2018 and 2019, which was followed by a precautionary advice since 2020. Since 2022 the stock is being assessed using an MSY approach; however, low SSB and the incapability of reaching Blim by 2024 even under with zero catches lead to a zero catch advice for 2023.

### 5.8 Future Issues and considerations

Cod in the Irish Sea and the Celtic Sea are in a highly exploited state and show historically a very steep age-profile. Recruitment since 2002 has been impeded.

It is essential to further the understanding of the stock structure to improve future management, which includes the further investigation of migration and natural mortality in the Irish Sea. It might be necessary for a combined approach to manage the stocks in 7.a and 7.e-g.

Under the current highly exploited status it seems that recruitment rather than fishing pressure is driving stock trends. It is also questionable in how far an MSY approach with reference points as applied in the traditional ICES format is a valid approach for this stock which is recruitment rather than fishery controlled. The working group is awaiting the outcomes of WKREF to further investigate the most appropriate way to manage the stock in the future. This might mean a shift to an MSE approach for management.

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Table 6. Official landings ( t ) of COD in Division 7.a as officially reported to ICES and figures used by ICES from 1996. All weights are in tonnes, minor differences in total value are due to rounding. Countries reported landings are official values.

|  |  | 凹 | O ¢ N O |  |  |  |  |  | $\stackrel{\text { ®0 }}{\square}$ |  |  | 出 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 142 | 148 | 2476 | 25 | - | 2359 | 27 | 126 | 5303 |  | 4964** |  |
| 1997 | 183 | 268 | 1492 | 29 | - | 2370 | 19 | 80 | 4441 |  | 5859** |  |
| 1998 | 316 | 269 | 1739 | 20 | - | 2517 | 34 | 67 | 4962 |  | 5318** |  |
| 1999 | 150 | $\mathrm{n} / \mathrm{a}$ | 966 | 5 | - | 1665 | 9 | 80 | 2875 |  | 4784** |  |
| 2000 | 60 | 53 | 455 | 1 | - | 799 | 11 | 38 | 1417 |  | 1274 |  |
| 2001 | 283 | 74 | 751 | - | - | 885 | 1 | 32 | 2026 |  | 2252 |  |
| 2002 | 318 | 116 | 1111 | - | - | 1134 | 7 | 29 | 2715 |  | 2695 |  |
| 2003 | 183 | 151 | 594 | - | 14 | 505 | 7 | 23 | 1477 |  | 1285 |  |
| 2004 | 104 | 29 | 380 | - | - | 646 | 5 | 15 | 1179 | 108 | 1072 |  |
| 2005 | 115 | 35 | 220 | - | - | 594 | n/a | 3 | 967 | 54 | 910 |  |
| 2006 | 60 | 18** | 275 | - | - | 589 | n/a | 6 | 948 | 103 | 840 |  |
| 2007 | 67 | 17** | 608 | - | - | 423 | n/a | 2 | 1117 | 527 | 702 | 148 |
| 2008 | 26 | 3 | 618** | - | - | 543 | 22 | 12 | 1224 | 558 | 661 | 62 |
| 2009 | 19 | 12 | 323** | - | - | 387 | 12 | 12 | 765 | 193 | 468 | 60 |
| 2010 | 21 | 1 | 289 | - | - | 282 | 1 | - | 594 | 143 | 464 | 377 |
| 2011 | 36 | 3 | 275 | - | - | 169 | 1 | - | 485 | 147 | 368 | 43 |


|  | $\stackrel{E}{\square}$ | U | O ¢ ¢ O |  | $\begin{aligned} & \text { 드주 } \\ & \text { in } \end{aligned}$ |  |  |  | $\stackrel{\bar{\square}}{\square}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 23 | 1 | 193 | - | - | 109 | <1 | - | 326 | 85 | 198 | 658 |
| 2013 | 13 | <1 | 160 |  |  | 107 | $<1$ | - | 281 | 76 | 206 | 118 |
| 2014 | 9 | <1 | 148 | - | - | 79 | $<1$ | - | 236 | 24 | 213 | 149 |
| 2015 | 12 | <1 | 137 | - | - | 50 | $<1$ | - | 199 | 39 | 161 | 224 |
| 2016 | 3 | <1 | 84 | - | - | 35 | $<1$ | - | 122 | 40 | 82 | 60 |
| 2017 | 5 | < 1 | 57 | - | - | 41 | <1 | <1 | 103 | 19 | 84 | 59 |
| 2018 | 2 | <1 | 105 | - | - | 128 | $<1$ | <1 | 235 | 20 | 215 | 42 |
| 2019 | 10 | < 1 | - C | - | - | 195 | <1 | <1 | 205 c | 37 | 295 | 7 |
| 2020* | 10 | 0 | 76 | - | - | 95 | <1 | <1 | 252 | 71 | 181 | 25 |
| 2021* | 3 | 0 | 93 | - | - | 89 | $<1$ | <1 | 184 | 52 | 133 | 4 |

* Preliminary official landings.
** Includes sample-based estimates of landings into ports.
*** Landings in the southern part of Division 7.a (rectangles 33 E 2 and 33E3) are not included in the assessment and are considered to be part of the cod stock in divisions 7.e-k.
c Incomplete/missing due to part of the data being unavailable under national GDPR clauses.

Table 7. Working Group figures for annual landings and TAC uptake since 2000. a) total, b) by country.
a)

| Year | Total | TAC | \% uptake |
| :---: | :---: | :---: | :---: |
| 2000 | 1273 | 2100 | 61 |
| 2001 | 2251 | 2100 | 107 |
| 2002 | 2695 | 3200 | 84 |
| 2003 | 1285 | 1950 | 66 |
| 2004 | 1072 | 2150 | 50 |
| 2005 | 910 | 2150 | 42 |
| 2006 | 840 | 1828 | 46 |
| 2007 | 702 | 1462 | 48 |
| 2008 | 662 | 1199 | 55 |
| 2009 | 468 | 899 | 52 |
| 2010 | 465 | 674 | 69 |
| 2011 | 368 | 506 | 73 |
| 2012 | 198 | 380 | 52 |
| 2013 | 206 | 285 | 72 |
| 2014 | 213 | 182 | 117 |
| 2015 | 161 | 146 | 110 |
| 2016 | 82 | 146 | 56 |
| 2017 | 84 | 146 | 57 |
| 2018 | 215 | 695 | 31 |
| 2019 | 298 | 807 | 37 |
| 2020 | 181 | 257 | 70 |
| 2021 | 133 | 206 | 65 |

b)

| 2009 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 391 | 55 | 3 | 19 | 0 | 498 |
| TAC | 259 | 592 | 33 | 12 | 3 | 899 |
| $\%$ uptake | $151 \%$ | $9 \%$ | $9 \%$ | $160 \%$ | $0 \%$ |  |


| 2010 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 292 | 151 | 1 | 21 | 0 | 465 |
| TAC | 194 | 444 | 25 | 9 | 2 | 674 |
| \% uptake | $150 \%$ | $34 \%$ | $4 \%$ | $233 \%$ | $0 \%$ |  |


| 2011 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 170 | 160 | 3 | 36 | 0 | 369 |
| TAC | 146 | 333 | 19 | 7 | 2 | 506 |
| \% uptake | $117 \%$ | $48 \%$ | $16 \%$ | $533 \%$ | $0 \%$ |  |


| 2012 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 112 | 63 | 0 | 23 | 0 | 198 |
| TAC | 109 | 251 | 14 | 5 | 1 | 380 |
| $\%$ uptake | $103 \%$ | $25 \%$ | $0 \%$ | $460 \%$ | $0 \%$ |  |


| 2013 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 107 | 85 | 1 | 13 | 0 | 206 |
| TAC | 82 | 188 | 10 | 4 | 1 | 285 |
| $\%$ uptake | $130 \%$ | $45 \%$ | $10 \%$ | $325 \%$ | $0 \%$ |  |


| 2014 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 79 | 124 | 0 | 9 | 0 | 213 |
| TAC | 52 | 120 | 7 | 2 | 2 | 182 |
| \% uptake | $153 \%$ | $103 \%$ | $0 \%$ | $455 \%$ | $0 \%$ |  |


| 2015 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 50 | 99 | 0 | 12 | 0 | 161 |
| TAC | 42 | 97 | 5 | 2 | 0 | 146 |
| \% uptake | $119 \%$ | $102 \%$ | $0 \%$ | $600 \%$ | NA |  |


| 2016 | UK | Ireland | France | Belgium | Netherlands | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 35 | 44 | 0.4 | 3 | 0 | 82 |
| TAC | 42 | 97 | 5 | 2 | 0 | 146 |
| \% uptake | 83\% | 45\% | 8\% | 150\% | 0\% |  |
| 2017 | UK | Ireland | France | Belgium | Netherlands | Total |
| Landings | 41 | 38 | 0.2 | 5 | 0 | 84 |
| TAC | 42 | 97 | 5 | 2 | 0 | 146 |
| \% uptake | 98\% | 39\% | 4\% | 250\% | 0\% |  |
| 2018 | UK | Ireland | France | Belgium | Netherlands | Total |
| Landings | 128.5 | 84.6 | 0.05 | 1.9 | 0 | 214.9 |
| TAC | 200 | 459 | 25 | 9 | 2 | 695 |
| \% uptake | 64\% | 18\% | <1\% | <1\% | 0\% | 31\% |


| 2019 | UK | Ireland | France | Belgium | Netherlands | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 193.9 | 90 | 0.2 | 10.2 | 0 | 294.6 |
| TAC | 233 | 530 | 30 | 11 | 3 | 807 |
| \% uptake | 83\% | 17\% | <1\% | 93\% | 0\% | 36.5\% |
| 2020 | UK | Ireland | France | Belgium | Netherlands | Total |
| Landings | 95.6 | 75.9 | 0 | 9.5 | 0 | 181.1 |
| TAC | 74 | 170 | 9 | 3 | 1 | 257 |
| \% uptake | 129\% | 45\% | 0\% | 317\% | 0\% | 70\% |
| 2021 | UK | Ireland | France | Belgium | Netherlands | Total |
| Landings | 88.7 | 41.8 | 0 | 2.8 | 0 | 133.3 |
| TAC | 91 | 104 | 7 | 3 | 1 | 206 |
| \%uptake | 97\% | 40\% | 0\% | 93\% | 0 | 65\% |

Table 8. Landings and discard proportions by métier.

| Catch (2021) | Landings |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 137 tonnes | otter trawls |  | midwater trawl | beam trawls | other gear types |
|  | Nephrops directed | demersal fish directed | 18.7\% | 9.4\% | 1\% |
|  | 33\% | 37\% |  |  |  |
|  | 133 tonnes |  |  |  |  |
|  | Discards |  |  |  |  |
|  | otter trawls |  | midwater trawl | beam trawls | other gear types |
|  | 77\% Nephrops directed | < $1 \%$ demersal fish directed | <1\% | 22\% | 1\% |
|  | 4 tonnes |  |  |  |  |

Table 9. Total catch numbers-at-age (thousands).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 17 | 439 | 1563 | 1003 | 456 | 177 | 30 |
| 1969 | 20 | 969 | 1481 | 1050 | 269 | 186 | 113 |
| 1970 | 22 | 1810 | 1385 | 352 | 204 | 163 | 71 |
| 1971 | 22 | 2835 | 2022 | 904 | 144 | 67 | 51 |
| 1972 | 26 | 900 | 3267 | 824 | 250 | 58 | 59 |
| 1973 | 27 | 2377 | 1091 | 1783 | 430 | 173 | 81 |
| 1974 | 16 | 601 | 3559 | 557 | 494 | 131 | 74 |
| 1975 | 26 | 1810 | 642 | 1407 | 294 | 249 | 117 |
| 1976 | 27 | 1247 | 3007 | 363 | 500 | 61 | 104 |
| 1977 | 31 | 946 | 511 | 1233 | 163 | 218 | 71 |
| 1978 | 40 | 855 | 1092 | 310 | 311 | 39 | 65 |
| 1979 | 44 | 1948 | 1288 | 608 | 127 | 164 | 71 |
| 1980 | 25 | 2636 | 2797 | 729 | 243 | 49 | 55 |
| 1981 | 38 | 1457 | 3635 | 1448 | 244 | 99 | 47 |
| 1982 | 46 | 538 | 2284 | 1455 | 557 | 102 | 79 |
| 1983 | 47 | 1011 | 932 | 751 | 499 | 154 | 46 |
| 1984 | 37 | 1733 | 1195 | 439 | 240 | 161 | 75 |
| 1985 | 34 | 1360 | 2105 | 703 | 158 | 84 | 77 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 49 | 1180 | 2248 | 699 | 203 | 64 | 65 |
| 1987 | 47 | 4522 | 1793 | 841 | 252 | 75 | 43 |
| 1988 | 43 | 2971 | 4734 | 702 | 263 | 71 | 38 |
| 1989 | 41 | 754 | 2163 | 1886 | 231 | 86 | 37 |
| 1990 | 38 | 869 | 1075 | 545 | 372 | 70 | 30 |
| 1991 | 47 | 2169 | 1408 | 442 | 127 | 98 | 22 |
| 1992 | 37 | 1529 | 1243 | 664 | 132 | 42 | 49 |
| 1993 | 39 | 388 | 2907 | 403 | 119 | 16 | 13 |
| 1994 | 40 | 916 | 569 | 848 | 68 | 20 | 10 |
| 1995 | 43 | 678 | 1283 | 180 | 163 | 7 | 6 |
| 1996 | 88 | 447 | 1113 | 700 | 38 | 39 | 6 |
| 1997 | 5 | 651 | 1149.5 | 501 | 213 | 17 | 16 |
| 1998 | 0 | 231 | 1928 | 335 | 80 | 28 | 8 |
| 1999 | 141 | 236 | 843 | 871 | 66 | 21 | 7 |
| 2000 | 62 | 1107 | 176 | 107 | 50 | 4 | 1 |
| 2001 | 7 | 403 | 841 | 53 | 13 | 9 | 2 |
| 2002 | 0 | 238 | 564 | 405 | 7 | 2 | 3 |
| 2003* | 50 | 121 | 472 | 109 | 36 | 1 | 0 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004* | 50 | 161 | 134 | 174 | 22 | 6 | 3 |
| 2005* | 50 | 118 | 256 | 78 | 34 | 5 | 1 |
| 2006 | 50 | 89 | 174 | 128 | 17 | 8 | 3 |
| 2007 | 16 | 216 | 210 | 56 | 11 | 1 | 0 |
| 2008 | 6 | 77 | 169 | 87 | 9 | 3 | 0 |
| 2009 | 329 | 60 | 57 | 66 | 17 | 3 | 0 |
| 2010 | 49 | 220 | 188 | 16 | 7.5 | 2 | 1 |
| 2011 | 10 | 54 | 106 | 36 | 2 | 1 | 1 |
| 2012 | 8 | 84 | 135 | 145 | 10 | 0 | 0 |
| 2013 | 36 | 37 | 59 | 30 | 9 | 2 | 0 |
| 2014 | 1 | 41 | 86 | 26 | 5 | 1 | 0 |
| 2015 | 0 | 37 | 80 | 26 | 4 | 1 | 0 |
| 2016 | 0 | 11 | 25 | 30 | 2 | 1 | 0 |
| 2017 | 0 | 12 | 28 | 16 | 3 | 0 | 0 |
| 2018 | 256 | 95 | 27 | 36 | 2 | 2 | 1 |
| 2019 | 0 | 60 | 68 | 12 | 9 | 1 | 2 |
| 2020* | 0 | 108 | 50 | 20 | 4 | 2 | 1 |
| 2021 | 0 | 11.8 | 22.1 | 13.1 | 4.7 | 0.3 | 0.7 |

*Excluded from assessment as very low sampling.

Table 10. Mean weights-at-age in the landings (used for whole stock and catch). *wean weight at age in landings only available for Q1, hence considerably lower than previous years and only used for forecast.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.86 |
| 1969 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.26 |
| 1970 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.17 |
| 1971 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.12 |
| 1972 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.28 |
| 1973 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.16 |
| 1974 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.34 |
| 1975 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.05 |
| 1976 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.13 |
| 1977 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.63 |
| 1978 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.19 |
| 1979 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.48 |
| 1980 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.87 |
| 1981 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.55 |
| 1982 | 0.1 | 1.01 | 1.52 | 3.49 | 5.57 | 7.59 | 9.11 |
| 1983 | 0.1 | 1 | 1.84 | 3.99 | 5.96 | 7.97 | 9.97 |
| 1984 | 0.1 | 0.68 | 1.81 | 3.81 | 5.87 | 7.48 | 10.05 |
| 1985 | 0.1 | 0.78 | 2.02 | 4.24 | 5.83 | 7.5 | 9.04 |
| 1986 | $0.1$ | 0.81 | $1.83$ | $3.86$ | 5.86 | $7.39$ | 8.78 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 0.1 | 0.71 | 2.16 | 3.91 | 6.41 | 7.82 | 10.32 |
| 1988 | 0.1 | 0.61 | 1.56 | 3.76 | 5.67 | 8.02 | 9.88 |
| 1989 | 0.1 | 0.94 | 1.85 | 3.22 | 5.41 | 6.57 | 9.47 |
| 1990 | 0.1 | 0.84 | 1.94 | 3.57 | 5.28 | 7.53 | 9.4 |
| 1991 | 0.1 | 0.86 | 1.64 | 3.54 | 5.42 | 6.39 | 9.11 |
| 1992 | 0.1 | 0.81 | 1.96 | 3.99 | 5.98 | 6.92 | 8.67 |
| 1993 | 0.1 | 0.85 | 1.71 | 3.67 | 5.68 | 7.37 | 10.17 |
| 1994 | 0.1 | 0.8 | 1.92 | 3.61 | 6.08 | 7.68 | 8.57 |
| 1995 | 0.1 | 0.9 | 1.84 | 4.00 | 5.79 | 8.45 | 9.14 |
| 1996 | 0.1 | 0.98 | 1.63 | 3.26 | 5.3 | 7.72 | 9.79 |
| 1997 | 0.1 | 0.85 | 1.94 | 3.62 | 5.29 | 6.12 | 9.4 |
| 1998 | 0.1 | 0.93 | 1.65 | 3.73 | 5.37 | 7.03 | 9.35 |
| 1999 | 0.1 | 0.85 | 1.62 | 3.18 | 5.51 | 7.52 | 10.25 |
| 2000 | 0.1 | 0.85 | 1.99 | 3.57 | 5.14 | 7.15 | 8.39 |
| 2001 | 0.1 | 0.99 | 1.82 | 4.15 | 5.61 | 7.33 | 9.51 |
| 2002 | 0.1 | 0.94 | 1.84 | 3.44 | 5.73 | 7.71 | 10.01 |
| 2003 | 0.1 | 1.21 | 1.66 | 3.29 | 5.43 | 10.2 | 11.09 |
| 2004 | 0.1 | 1.11 | 2.2 | 3.63 | 6.51 | 7.64 | 8.61 |
| 2005 | 0.1 | 0.91 | 1.94 | 3.51 | 5.32 | 7.74 | 8.89 |
| 2006 | 0.1 | 0.83 | 1.84 | 3.67 | 4.71 | $6.39$ | 7.84 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 0.1 | 0.83 | 1.85 | 3.78 | 5.35 | 7.99 | 10.04 |
| 2008 | 0.1 | 0.89 | 1.59 | 3.54 | 6.00 | 7.57 | 9.46 |
| 2009 | 0.1 | 1.1 | 2.01 | 3.46 | 5.31 | 7.1 | 6.82 |
| 2010 | 0.1 | 1.26 | 2.29 | 3.93 | 6.34 | 7.33 | 9.64 |
| 2011 | 0.1 | 0.95 | 1.88 | 3.75 | 5.54 | 6.75 | 9.04 |
| 2012 | 0.1 | 0.93 | 1.88 | 3.37 | 5.34 | 7.60 | 8.56 |
| 2013 | 0.1 | 0.97 | 2.32 | 4.06 | 5.54 | 7.43 | 10.79 |
| 2014 | 0.1 | 0.88 | 2.26 | 4.49 | 7.00 | 8.75 | 9.41 |
| 2015 | 0.1 | 0.83 | 1.79 | 3.69 | 6.49 | 8.55 | 9.95 |
| 2016 | 0.1 | 0.95 | 1.58 | 3.1 | 5.01 | 10.66 | 8.136 |
| 2017 | 0.1 | 0.70 | 1.82 | 3.82 | 5.85 | 7.62 | 9.74 |
| 2018 | 0.1 | 0.43 | 1.69 | 3.64 | 5.56 | 8.58 | 8.70 |
| 2019 | NA | 0.44 | 2.13 | 4.25 | 6.14 | 6.79 | 9.00 |
| 2020 * | 0.1 | 0.22 | 1.29 | 3.67 | 5.23 | 7.85 | 9.54 |
| 2021 | 0.1 | 0.187 | 1.831 | 4.164 | 6.485 | 8.64 | 7.25 |

Table 11. Estimates of numbers discarded (a) and the discarded proportions (b) from 1968-2021. Data are total numbers ('000 fish) discarded at-age, estimated from numbers per sampled trip raised to total fishing effort by each country supplying data (UK, Ireland and Belgium) Please refer to WKIrish3 (ICES, 2017a) documents.
a)

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 17.81 | 74.71 | 0 | 0 | 0 | 0 | 0 |
| 1969 | 20.85 | 87.45 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 22.13 | 92.83 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 22.94 | 96.2 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 26.51 | 111.18 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 27.17 | 113.96 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 16.94 | 71.04 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 26.38 | 110.62 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 26.77 | 112.28 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 31.05 | 130.23 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 39.96 | 167.57 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 44.35 | 185.98 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 24.6 | 103.16 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 37.67 | 157.97 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 46.04 | 193.1 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 46.98 | 197.05 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 37.3 | 156.45 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 33.89 | 142.12 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 49.15 | 206.15 | 0 | 0 | 0 | 0 | 0 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 47.38 | 198.69 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 42.59 | 178.64 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 41.03 | 172.09 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 37.85 | 158.74 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 46.64 | 195.61 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 36.74 | 154.1 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 39.4 | 165.24 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 39.92 | 167.44 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 42.97 | 180.2 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 87.95 | 128.79 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 5.28 | 127.79 | 0.5 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 27.47 | 2 | 0 | 0 | 0 | 0 |
| 1999 | 141.42 | 165.79 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 62.36 | 817.69 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 7.22 | 65.15 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 42.49 | 0 | 0 | 0 | 0 | 0 |
| 2003 * | 50.43 | 75.68 | 32.62 | 15.83 | 1.25 | 0.13 | 0 |
| 2004* | 50.43 | 92.78 | 32.81 | 15.83 | 1.25 | 0.13 | 0 |
| 2005* | 50.43 | 76.34 | 32.36 | 15.83 | 1.25 | 0.13 | 0 |
| 2006 | 50.43 | 75.08 | 32 | 15.83 | 1.25 | 0.13 | 0 |
| 2007 | 16 | 167 | $4.60$ | 0 | 0 | 0 | 0 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 5.50 | 63.40 | 3.40 | 0 | 0 | 0 | 0 |
| 2009 | 329.30 | 39.80 | 4.40 | 0.1 | 0 | 0 | 0 |
| 2010 | 48.70 | 180 | 60.30 | 1.4 | 0.5 | 0.1 | 0 |
| 2011 | 9.70 | 42.70 | 0.90 | 0 | 0 | 0 | 0 |
| 2012 | 7.50 | 79.90 | 100.20 | 112.9 | 5.9 | 0.2 | 0 |
| 2013 | 36.10 | 31 | 26.50 | 11 | 2 | 0.5 | 0 |
| 2014 | 1.09 | 34.66 | 41.93 | 10.3 | 1.53 | 0.1 | 0 |
| 2015 | 0 | 37.30 | 45.80 | 6.8 | 1.3 | 0.3 | 0 |
| 2016 | 0 | 9.84 | 14.15 | 13.45 | 0.91 | 0.74 | 0 |
| 2017 | 0.43 | 9.85 | 7.88 | 8.10 | 0.57 | 0.10 | 0.10 |
| 2018 | 255.50 | 72.19 | 8.89 | 4.88 | 0.12 | 0.22 | 0 |
| 2019 | 0 | 39.2 | 0.4 | 0 | 0 | 0 | 0 |
| 2020* | NA |  |  |  |  |  |  |
| 2021 | 0 | $10.6$ | $6.1$ | $0$ | $0$ | $0$ | 0 |

* very low sampling levels.
b)

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 1 | 0.17 | 0 | 0 | 0 | 0 | 0 |
| 1969 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 1 | 0.05 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 1 | 0.03 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 1 | 0.12 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 1 | 0.05 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 1 | 0.12 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 1 | 0.06 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 1 | 0.14 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 1 | 0.20 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 1 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 1 | 0.04 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 1 | 0.11 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 1 | 0.36 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 1 | 0.19 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 1 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 1 | 0.17 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 1 | 0.04 | $0$ | $0$ | $0$ | $0$ | $0$ |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 1 | 0.06 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 1 | 0.23 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 1 | 0.18 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 1 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 1 | 0.43 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 1 | 0.18 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 1 | 0.27 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 1 | 0.29 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 1 | 0.20 | 0 | 0 | 0 | 0 | 0 |
| 1998 | NA | 0.12 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 1 | 0.70 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 1 | 0.74 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 1 | 0.16 | 0 | 0 | 0 | 0 | 0 |
| 2002 | NA | 0.18 | 0 | 0 | 0 | 0 | 0 |
| 2003 * | 1 | 0.63 | 0.07 | 0.15 | 0.03 | 0.12 | NA |
| 2004* | 1 | 0.58 | 0.25 | 0.09 | 0.06 | 0.022 | 0 |
| 2005* | 1 | 0.65 | 0.13 | 0.20 | 0.04 | 0.03 | 0 |
| 2006 | 1 | 0.84 | 0.18 | 0.12 | 0.07 | 0.02 | 0 |
| 2007 | 1 | 0.77 | 0.02 | 0 | 0 | 0 | NA |
| 2008 | $1$ | 0.82 | $0.02$ | $0$ | $0$ |  | NA |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 1 | 0.67 | 0.08 | 0 | 0 | 0 | NA |
| 2010 | 1 | 0.82 | 0.32 | 0.06 | 0.07 | 0.05 | 0 |
| 2011 | 1 | 0.80 | 0.01 | 0 | 0 | 0 | 0 |
| 2012 | 1 | 0.95 | 0.74 | 0.78 | 0.60 | 1 | NA |
| 2013 | 1 | 0.84 | 0.45 | 0.37 | 0.22 | 0.34 | NA |
| 2014 | 1 | 0.85 | 0.49 | 0.39 | 0.28 | 0.09 | NA |
| 2015 | NA | 1 | 0.57 | 0.26 | 0.30 | 0.23 | NA |
| 2016 | NA | 0.91 | 0.58 | 0.45 | 0.40 | 0.62 | 0 |
| 2017 | 1 | 0.80 | 0.28 | 0.51 | 0.20 | 0.21 | 0.49 |
| 2018 | 1 | 0.76 | 0.33 | 0.13 | 0.05 | 0.10 | 0 |
| 2019 | NA | 0.65 | <0.01 | 0 | 0 | 0 | 0 |
| 2020* |  |  |  |  |  |  |  |
| 2021 | 1 | 0.89 | $0.28$ | $0$ | $0$ | $0$ | 0 |

NA= not available.
${ }^{*}$ Data for are unavailable due to restricted discard sampling.

| Year | 1 | 2 | 3+ |
| :---: | :---: | :---: | :---: |
| 1996 | 0 | 0.27 | 1 |
| 1997 | 0 | 0.275415 | 1 |
| 1998 | 0 | 0.339514 | 1 |
| 1999 | 0 | 0.402555 | 1 |
| 2000 | 0 | 0.464725 | 1 |
| 2001 | 0 | 0.526111 | 1 |
| 2002 | 0 | 0.585231 | 1 |
| 2003 | 0 | 0.623356 | 1 |
| 2004 | 0 | 0.65373 | 1 |
| 2005 | 0 | 0.676757 | 1 |
| 2006 | 0 | 0.691103 | 1 |
| 2007 | 0 | 0.697111 | 1 |
| 2008 | 0 | 0.700228 | 1 |
| 2009 | 0 | 0.704985 | 1 |
| 2010 | 0 | 0.707035 | 1 |
| 2011 | 0 | 0.704413 | 1 |
| 2012 | 0 | 0.700372 | 1 |
| 2013 | 0 | 0.702394 | 1 |


| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3 +}$ |
| :--- | :--- | :--- | :--- | :--- |
| 2014 | 0 | 0.708485 | 1 |
| 2015 | 0 | 0.716712 | 1 |
| 2016 | 0 | 0.726138 | 1 |
| 2017 | 0 | 0.735987 | 1 |
| 2018 | 0 | 0.745951 | 1 |
| 2019 | 0 | 0.756372 | 1 |
| 2020 | 0 | 0.74887 | 1 |
| 2021 | 0 | 0.75601 | 1 |

## Table 13. Survey catch numbers-at-age and c.v. for all three surveys.

Survey catch numbers-at-age and c.v.

| year | c.v. | 1 | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 0.68 | 700.73 | 386.15 | 20.03 | 10.78 |
| 1996 | 0.42 | 1106.13 | 329.28 | 111.67 | 1.39 |
| 1997 | 0.64 | 537.30 | 415.84 | 66.72 | 21.39 |
| 1998 | 0.84 | 169.36 | 769.23 | 56.87 | 11.98 |
| 1999 | 0.86 | 49.50 | 253.08 | 241.87 | 15.29 |
| 2000 | 0.65 | 629.60 | 101.053 | 34.58 | 33.01 |
| 2001 | 0.89 | 406.68 | 561.44 | 18.44 | 5.78 |
| 2002 | 0.64 | 662.16 | 253.31 | 333.54 | 0 |
| 2003 | 0.54 | 73.87 | 1079.20 | 104.05 | 32.70 |
| 2004 | 0.75 | 216.96 | 171.96 | 88.62 | 5.38 |
| 2005 | 0.76 | 63.53 | 225.07 | 29.41 | 27.96 |
| 2006 | 0.63 | 169.99 | 130.75 | 58.30 | 2.52 |
| 2007 | 0.95 | 164.35 | 124.39 | 30.60 | 5.15 |
| 2008 | 0.90 | 40.66 | 217.15 | 13.02 | 5.17 |
| 2009 | 0.76 | 144.00 | 59.00 | 33.00 | 9.00 |
| 2010 | 0.82 | 1022.12 | 208.96 | 14.66 | 2.26 |
| 2011 | 0.49 | 353.98 | 414.69 | 46.01 | 2.26 |


| year | c.v. | 1 | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 0.81 | 161.90 | 222.82 | 99.27 | 14.25 |
| 2013 | 0.81 | 276.59 | 213.68 | 60.08 | 1.49 |
| 2014 | 0.63 | 314.41 | 222.80 | 53.29 | 13.66 |
| 2015 | 0.84 | 78.96 | 719.35 | 69.19 | 8.56 |
| 2016 | 1.06 | 349.20 | 175.00 | 148.30 | 10.70 |
| 2017 | 0.77 | 69.8 | 445.20 | 57.80 | 12.60 |
| 2018 | 1.26 | 138.1 | 50.50 | 62.60 | 0 |
| 2019 | 0.88 | 214.9 | 171.6 | 27.8 | 14.7 |
| 2020 | 0.977 | 78.5 | 145.4 | 39.4 | 0 |
| 2021 | 1.19 | 86.1 | 158.9 | 38.2 | 0 |

Northern Irish Groundfish Quarter 4

| year | c.v. | 0 |
| :---: | :---: | :---: |
| 1995 | 0.54163 | 6.66 |
| 1996 | 0.430336 | 12.519 |
| 1997 | 0.720571 | 2.345 |
| 1998 | 0.914513 | 0.047 |
| 1999 | 0.637233 | 6.734 |
| 2000 | 0.785349 | 6.212 |
| 2001 | 0.830289 | 4.863 |
| 2002 | 0.895678 | 0.123 |
| 2003 | 0.707142 | 6.746 |
| 2004 | 0.939137 | 3.663 |
| 2005 | 0.805428 | 8.144 |
| 2006 | 0.871324 | 1.16 |
| 2007 | 1.277817 | 0.067 |
| 2008 | 1.422627 | 0.185 |
| 2009 | 0.938364 | 5.356 |
| 2010 | 1.332794 | 2.779 |
| 2011 | 0.919446 | 0.084 |
| 2012 | 1.256171 | 1.924 |
| 2013 | 0.933411 | 11.208 |
| 2014 | 0.792604 | 0.121 |
| 2015 | 0.872952 | 2.244 |
| 2016 | 1.063181 | 0.149 |
| 2017 | 0.815541 | 4.291 |
| 2018 | 1.419523 | 0.685 |
| 2019 | 1.266571 | 0.072 |
| 2020 | 1.386682 | 0.072 |
| 2021 | 1.610235 | 0.335 |

UK FSP survey

| year | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0.43 | 1.41 | 0.99 | 0.08 | 0.03 |
| 2006 | 0.54 | 2.81 | 0.43 | 0.10 | 0.01 |
| 2007 | 0.61 | 1.32 | 0.59 | 0.06 | 0.06 |
| 2008 | 0.22 | 0.82 | 0.15 | 0.08 | 0.02 |
| 2009 | 0.17 | 1.15 | 0.38 | 0.10 | 0.02 |
| 2010 | 0.74 | 0.45 | 0.47 | 0.13 | 0.02 |
| 2011 | 0.41 | 1.68 | 0.14 | 0.10 | 0.04 |
| 2012 | 0.36 | 2.30 | 0.80 | 0.07 | 0.02 |
| 2013 | 0.84 | 1.88 | 1.35 | 0.37 | 0.06 |
| 2014 |  |  |  |  |  |
| 2015 | 0.60 | 2.04 | 1.17 | 0.26 | 0.05 |
| 2016 | 1.00 | 6.39 | 1.43 | 0.41 | 0.03 |
| 2017 | 3.06 | 2.85 | 3.84 | 1.01 | 0.23 |
| 2018 | 0.43 | 3.73 | 0.61 | 0.63 | 0.15 |
| 2019 | 1.30 | 0.75 | 0.83 | 0.12 | 0.19 |
| 2020 | 0.77 | 2.64 | 0.13 | 0.18 | 0.08 |
| 2021 | 0.24 | 0.71 | 0.19 | 0.01 | 0.027 |

Q1 groundfish survey CPUE and SD used in the assessment.

| Year | CPUE | SD |
| :---: | :---: | :---: |
| 1995 | 0.955344 | 0.214285 |
| 1996 | 1.728974 | 0.313405 |
| 1997 | 1.391875 | 0.217769 |
| 1998 | 1.435543 | 0.198929 |
| 1999 | 1.597456 | 0.255936 |
| 2000 | 1.023321 | 0.146161 |
| 2001 | 1.491194 | 0.224681 |
| 2002 | 2.619399 | 0.964573 |
| 2003 | 1.696543 | 0.235312 |
| 2004 | 0.764752 | 0.139312 |
| 2005 | 0.890243 | 0.267329 |
| 2006 | 0.508091 | 0.07914 |
| 2007 | 0.46498 | 0.104631 |
| 2008 | 0.501744 | 0.098636 |
| 2009 | 0.494051 | 0.141257 |
| 2010 | 0.71933 | 0.129658 |
| 2011 | 1.204889 | 0.364965 |
| 2012 | 1.017556 | 0.179033 |
| 2013 | 1.074564 | 0.205801 |
| 2014 | 1.089111 | 0.274391 |
| 2015 | 1.785167 | 0.26655 |
| 2016 | 1.374257 | 0.246976 |
| 2017 | 1.029783 | 0.30429 |
| 2018 | 0.631522 | 0.11959 |
| 2019 | 0.816597 | 0.221725 |
| 2020 | 0.492889 | 0.177333 |
| 2021 | 0.476304 | 0.122131 |

## Table 14. Assessment summary.

| Year | Recruitment | SSB |  |  |  |  | Landings | Discards | Fishing mortality ages 2-4 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Value | High | Low | Value | High | Low |  |  | Value | High | Low |
| 1968 | 184549 | 250405 | 118693 | 46341 | 57748 | 34934 | 8541 | 1285 | 0.108 | 0.156 | 0.059 |
| 1969 | 248327 | 326531 | 170123 | 41512 | 52782 | 30242 | 7991 | 1898 | 0.26 | 0.31 | 0.196 |
| 1970 | 384744 | 489996 | 279492 | 36819 | 47717 | 25921 | 6426 | 708 | 0.28 | 0.33 | 0.24 |
| 1971 | 139637 | 192753 | 86521 | 36214 | 47328 | 25101 | 9246 | 363 | 0.22 | 0.28 | 0.162 |
| 1972 | 358296 | 451004 | 265588 | 41664 | 54267 | 29061 | 9234 | 1546 | 0.26 | 0.33 | 0.194 |
| 1973 | 89533 | 127165 | 51901 | 46927 | 61667 | 32188 | 11819 | 1222 | 0.26 | 0.35 | 0.166 |
| 1974 | 286027 | 363839 | 208215 | 39749 | 52783 | 26715 | 10251 | 1749 | 0.34 | 0.42 | 0.25 |
| 1975 | 95891 | 133523 | 58259 | 40510 | 53818 | 27201 | 9863 | 857 | 0.32 | 0.42 | 0.23 |
| 1976 | 152041 | 202217 | 101865 | 31895 | 43028 | 20762 | 10247 | 381 | 0.33 | 0.44 | 0.23 |
| 1977 | 156311 | 207075 | 105547 | 31257 | 42762 | 19752 | 8054 | 201 | 0.36 | 0.46 | 0.25 |
| 1978 | 292095 | 365791 | 218399 | 24799 | 34579 | 15019 | 5662 | 0 | 0.33 | 0.40 | 0.25 |
| 1979 | 325000 | 401048 | 248952 | 24541 | 33713 | 15368 | 7548 | 0 | 0.25 | 0.34 | 0.162 |
| 1980 | 183575 | 236103 | 131047 | 28154 | 36916 | 19393 | 10599 | 0 | 0.30 | 0.39 | 0.22 |
| 1981 | 87444 | 119000 | 55888 | 36538 | 46259 | 26816 | 13958 | 0 | 0.35 | 0.44 | 0.25 |




| Year | Recruitment |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| age 0 |  |
| Value |  |

*Geometric Mean 2002 to 2019.


Figure 1. Available data.


Figure 2. Landings and discards-at-age. Landings are shaded in grey, discards in white.


Figure 3. Proportion discarded-at-age. Ages 1 and 0 not displayed.


Figure 4. SSB with 95\% confidence interval.


Figure 5. Recruitment with $95 \%$ confidence level. Recruitment in the figure for 2022 is model estimated and not the same as in the forecast.


Figure 6. Age compositions for commercial data and surveys.


Figure 7. Residuals at-age.


Figure 8. Log CPUE fit NIGFS Q1.


Figure 9. Log index fit NIGFS Q4.


Figure 10. Log index fit UKFSP survey.


Figure 11. Results for runs tests for the three indices included and RMSE with fitted LOESS smoother.


Figure 12. Mean age residual fits for total catches, NIGFSQ1 and UKFSP surveys, NIGFSQ4 survey only includes age 0 recruits and is therefore excluded.


Figure 13. Mohns Rho for SSB and $\mathrm{F}_{\text {bar }}$.

## 6 Cod in divisions 7.e-k (eastern English Channel and southern Celtic Seas)

## Full analytical assessment

This stock has been benchmarked at WKCELTIC 2020. XSA was replace by SAM as the assessment model. Time-series of data were updated since 2004 as well as the tuning series. The first ten years of data (1970-1979) were removed from the assessment time-series of catches, because of inconsistency in cohort tracking information. Data, assessment and forecast procedure are detailed in the stock annex.

## Latest ICES advices in 2020 and 2021

2020 - "For Cod in divisions 7.e-k, ICES advises that when the MSY approach is applied, there should be zero catch in 2021."

2021 - "For Cod in divisions 7.e-k, ICES advises that when the MSY approach and precautionary considerations are applied, there should be zero catch in 2022."

### 6.1 General

6.2 Stock description and management units

The TAC is set for ICES areas $7 . \mathrm{b}-\mathrm{c}, 7 . \mathrm{e}-\mathrm{k}, 8,10$, and CECAF 34.1.1(1), excluding 7.d. This is representative of the stock area as the cod population in $7 . \mathrm{d}$ is more relevant to the North Sea population. However, landings from 7.bc are not included in the assessment area.

## Management applicable in 2021 and 2022

TAC 2021 (Council regulation 2021/1239)

| Species: | Cod Gadus morhua |  | Zone: | 7b, 7c, 7e-k, 8, 9 and 10; Union waters of CECAF <br> 34.1.1 <br> (COD/7XAD 34) |
| :---: | :---: | :---: | :---: | :---: |
| Belgium | 18 | a) | Analyt |  |
| France | 290 | (1) | Article | is Regulation applies |
| Ireland | 422 | (1) | Article | gulation (EC) No 847/96 shall not apply |
| The <br> Netherlands | 0 | (a) | Article | gulation (EC) No 847/96 shall not apply |
| Union | 730 | a) |  |  |
| United Kingdom | 75 | a) |  |  |
| TAC | 805 |  |  |  |
| (1) | Exclusively for by-catches of cod in fisheries for other species. No directed fisheries for cod are permitted under this quota. |  |  |  |

## Preliminary TAC 2022 (Council regulation 2022/109)



Since 2005, ICES rectangles 30E4, 31E4, and 32E3 have been closed during the first quarter (Council Regulations 27/2005, 51/2006, and 41/2007, 40/2008, and 43/2009).

Technical measures applied to this stock are a minimum mesh size (MMS) for beam and otter trawlers in Subarea 7 and a minimum landing size (MLS) of 35 cm .

## Fishery

Landings data used by the WG are summarised in Table 6.1 and the Figure 6.1 provides historical landings by countries. In 2021, the catches are 1360 t .

TAC was overtaken in 2021. An overtaking of the agreed TAC was observed for France, while Belgium, Ireland and UK were closed to the TAC. Cod is no longer a target species but a bycatch in haddock and whiting dedicated fisheries.

Given the rapid growth of cod in this area, discards are mostly composed of one and two yearold fish. Since 2011, quotas were not restricted and the discard rate has been stable around 10$15 \%$. However, following the recent TAC reductions, TAC is now restrictive for most of the countries. Discards estimate for 2021 is 733 t . It corresponds to a discards rate of $54 \%$, which is significantly greater than the average discards rate of recent years (around 20\%). This discards rate increase may be the result of high grading, because of restrictive TACs and delay in total TAC attribution. This delay was mainly due to long discussions in Brexit fisheries negotiations.

Cod is mainly caught in area 27.7.g, followed by areas 27.7.h, 27.7.e and 27.7.j respectively. No landings are reported in 27.7.k and few in 27.7.j2 (Figure 6.2). France is fishing in all areas but most of its landings are taking in 27.7.h. Ireland is mainly fishing in 27.7.g and Belgium in 27.7.f and UK in 27.7.e. For each country, landings distribution in the Celtic Sea is similar to previous years.

In Celtic Sea, cod is mainly caught by OTB_DEF_100-119_0_0_all métiers ( $37 \%$ of the landings and $75 \%$ of the catches), followed by OTB_DEF_70-99_0_0_all, OTB_CRU_100-119_0_0_all and seine SCC_DEF_100-119_0_0_all. Beamers (i.e. TBB_DEF_70-99_0_0_all) also contribute to cod landings (Figure 6.3).

Discards rate in weight varies among métiers depending on gear, mesh size range and targeted species (Figure 6.4).

The group advises to follow métier definition specified in the Appendix 2 of the ICES data call to reduce the number of métier upload in InterCatch. Métier which contribute to less than $1 \%$ of the landings should be included in the MIS_MIS_0_0_0_HC métier.

## Information from the industry

In recent years, yields have been very low and cod is no longer targeted by French vessels and catches represent a very low number of individuals per tow.
The recent regulatory changes in the Celtic Sea since 2019 (Reg UE 2034/2018 which introduces many new selective devices since 01/07/2019 and article 13 Reg UE 123/2020) significantly modifies (1) the size structure of species catches by improving selectivity and the (2) vessel strategy in order to respect different catch composition thresholds.

### 6.3 Data

## InterCatch procedure

Since 2013, international landings and discards data are uploaded in InterCatch. An updated data tile series, from 2004 to 2019, was provided as part of the WKCELTIC 2020. Discards are raised for unreported strata to estimate total discards in weight. During WKCELTIC efforts were made to streamline data compilation procedures for fishery-dependent data of the three main gadoids species (cod, haddock and whiting).

Unsampled strata of landings and discards (number-at-age) are filled in using an allocation procedure. Information on national and international assumptions made by data providers and submitters at the national level and allocation grouping used in IC are available on SharePoint ( R script). To ensure the consistency of data processing at international level, the same rules are applied each year for the allocation procedure: fill unsampled strata using as much as possible the same métier and quarter, regardless of area and country. Unsampled BMS landings and Logbook Registered Discards are filled in using discards data employing as much as possible the same métier and quarter, regardless of area and country.

The impact of the Covid-19 pandemic on the fishery cannot be quantitatively determined but may be assumed to have reduced fishing effort in quarter 2 of 2020.

The percentage of sampled versus raised data as well as the distribution of sampled data over the quarters were considered satisfactory (Figure 6.5).

| Season | Source | $\%$ |  |
| :--- | :--- | :--- | :--- |
|  | 1 | Imported | 23 |
| 2 | Imported | 45 |  |
| 3 | Imported | 11 |  |
|  | Imported | 19 |  |

## Catches

Age distribution of 2021 catches (i.e. landings and discards) is illustrated in the Figure 6.10 and Table 6.2. It is noticeable that this stock has always been composed of few age classes, even though Celtic Sea cod can live up to ten years. While the catch was mainly composed of age 2 over the period 2005-2008, the strong 2009 year class has contributed strongly to the catch at older ages in recent years: $63 \%$ in number in 2012 at age 3, $36 \%$ at age 4 in 2013 (Table 8.2.a and 8.2.b). In 2014, high recruitment has been observed resulting in an increasing proportion of age 1 fish in the landings ( $53 \%$ ), age 2 accounts for $22 \%$ of the landings. In 2015 , landings are dominated by fish of age 2, in 2016 landings are dominated by fish of age 3 and in 2017 landings was mostly composed by cod of age 2 . In 2018, $20 \%$ of the landings was fish of age $1,35 \%$ of fish of age 2 and $31 \%$ of fish of age 3. In 2019, more than $50 \%$ of the catches are age 1 fish, and less than $30 \%$ of the catches are made of age superior to 2 . In $2020,36 \%$ of the catches are age 1 fish, and $58 \%$ of the catches (in number) are made of age 2. In 2021, age 1 and age 2 represent each $40 \%$ of the catches (in number).

## Discards

The landings/discards pattern is known to be strongly variable between fleets and years due to métier, recruitment intensity, TACs constraints and mixed fisheries concerns.

In 2009, age 1 individuals ( $30-45 \mathrm{~cm}$, Mahé et al., 2016) were mainly discarded. In 2010, most of them were landed. In 2011, ages 1 and 2 represents respectively $51 \%$ and $46 \%$ of the total discards in numbers for all fleets. Due to the low TAC relative to the high magnitude of recruitment in 2009 and 2010, all countries had unusually high discard rates in 2011, generally $70 \%$ by weight was made up of fish above the minimum landing size (MLS, i.e. 35 cm for Celtic Sea cod). The high-graded fish from the French fishery have been added to the landings in 2003-2011. In 2014, total amount of discards was 740 t ( 639 t imported +101 t raised), giving a discard rate of $19 \%$. This discards rate was higher than the average $10 \%$ and mostly consisted of undersized fish from the strong 2013 year class (fish of age 1 in 2014). In 2015, the total amount of discard was 565 t ( 250 t sampled and uploaded in InterCatch and 309 t resulting from the raising procedures), giving a discard rate by weight of $12 \%$, which is considered the usual discard rate for this species in the mixed fisheries. High grading in 2015 (discards of fish above Minimum conservation size) was low. In 2016, the total amount of discards was 220 t ( 154 t sampled and uploaded in InterCatch and 52 t resulting from the raising procedures), giving a discard rate by weight of $6.3 \%$. In 2017, the total amount of discards was $117 \mathrm{t}(47 \mathrm{t}$ sampled and uploaded in InterCatch and 62 t resulting from the raising procedures), giving a discard rate by weight of $5 \%$, which is considered lower than average. They are mainly composed of age 1 fish (Figure 6.10).

In recent years, due to quota constraints at vessels levels, length distribution of discards for the UK fleet have shown high-grading pattern (cod being a non-target species). However, this fleet has little contribution to both, landings and discards quantities and this was no more reported in 2017. In 2019, discards are mostly composed of fish of 1 year, as in 2018 (Figure 6.10).

In 2021, French fleet have recorded high-grading pattern for its discards, maybe due to restrictive TACs and delay in total TAC attribution. Individual TAC allocations were attributed in three times and the entire individual TAC was allocated in June. The two preliminary individual TAC, allocated for the period between January and March and between March and July, were reached before the end of their corresponding periods which have led to discard high grading in Q1 and 2.


## Biological

Catch numbers-at-age, catch weights-at-age and stock weights-at-age are given respectively in Tables 6.2, 6.3 and 6.4.

Temporal trends in stock and catch were scrutinized at WGCSE 2021, to ensure that reduce sampling due to Covid-19 pandemic did not impact catch weight. No important issues were reported.


Biological parameters are described in the stock annex which has been updated at WKCELTIC 2020. Celtic Sea cod are very fast growing and early maturing compared with more northern cod stocks.

## Commercial LPUE

Tables 6.5 a-c gather the values of landings, fishing effort and LPUE dataseries for the French (a), Irish (b) and UK fleets (c). Figures 6.6 a-c illustrate the trends of LPUE and effort by country.

A general decrease in the LPUE trend is observed in almost all series between 1990 and 2004, where the TAC began to be constraining. From that point, the LPUE seemed to stabilize, or even to increase if high grading is taken into account. The strong 2009 year class resulted in an increase of LPUE for all fleets between 2010 and 2012. Different features are observed in the effort timeseries. The métiers showing the highest levels of cod directed effort have decreased significantly in the last 5-10 years until 2010. Since then, effort has gone up again until 2013 following the increased of TAC possibilities.

Since 2013, French fishing effort and LPUE have decreased (Figure 6.6a). Effort of Irish fleet targeting gadoids (i.e. Otter trawl 27.7.g) remains at a high level as a consequence of mixed-fisheries interaction with increased whiting and haddock fisheries opportunities (Figure 6.6b). In the meantime, the Spawning-Stock Biomass (SSB) is low, as such LPUE is decreasing since 2013. In 2018, Otter trawl Irish 27.7.g LPUE has increased. Effort of the UK trawl fleet in 27.7.e-k shows a decreasing trend (down to zero in 2016) and increases since then, while beam trawl effort in 27.7.e-k relatively stable in recent years (Figure 6.6c). Minor revision of FRA commercial fleet from 0.84 to 1.18 in 2019 was made, which is believed to have very little impact on assessment results.


The impact of the Covid-19 pandemic on the fishery cannot be quantitatively determined, but a slight reduction of fishing effort of the main fleets in 2020 was observed for all country. As, a result in 2020, LPUE of Irish otter trawls in 7g and UK trawls in 7ek are decreasing, while French otter trawl LPUE remain stable.

Remark: The UK English and Welsh effort data are only reliable for vessels over 12 metres registered length, and therefore has always been provided to working groups for vessels greater than 12 metres. The fleet of vessels over 12 meter has been declining gradually over the years, until in 2016 no effort recorded from this fleet. The zero figures provided for 2016 have been checked and are correct (Figure 6.6c).

## Surveys and commercial tuning fleet

Two ongoing surveys, both part of the DCF, IBTS Q4 (EVHOE-WIBTS-Q4; IGFS-WIBTS-Q4) are combined and modelled to produce a single index using VAST modelling (see details in the stock annex and WKCELTIC 2020 report).

In 2017 and 2018, the French EVHOE survey was not conducted due to technical difficulties at the beginning of the survey. The Irish survey covered additional stations normally undertaken by the EVHOE survey.

Commercial tuning index based on French OTB and OTT fleet is provided. The calculation of the commercial tuning series was updated at WKCELIC 2020 to better account for changes in fleet behaviour along the years (see details in the stock annex and WKCELTIC 2020 report). LPUE is decreasing since 2012.

The historical time-series of commercial tuning index (OTDEF French fleet for quarters 2, 3 and $4)$, and the survey index are shown in Table 6.6.

## Data issues

No important issues were reported this year. Owing to notable divergence of recent discards patterns, it was not possible to forecast separate landings and discards estimates for 2023.

Catch sampling of the fisheries has been reduced in 2020 due to Covid-19, which may have result in a higher uncertainty associated with discard estimates and age structure of the catch. However, this was considered to have had minimal impact on the perception of the stock status.

Remark: When for a métier/strata landings are upload annually, there are no information available in InterCatch to split the annual landings into quarterly landings and therefore the associated age composition and mean weight-at-age. As a result, when extracting quarter 1 versus
quarters 2,3 and, 4 data to inform on mean weight of the stock and the catch for the assessment, these data are not used.

### 6.4 Stock assessment

Model used: SAM (stockassessment.org).

## Final update assessment (SAM)

The final assessment was run with the same settings as established by WKCELTIC 2020 and described in the stock annex. Discards are included in the assessment. (sotcokassessment.org, Cod_7ek_WGCSE2020).

Residuals and diagnostics do not highlight any problem regarding the input data and model fit (Figure 6.7 and 6.8). Outputs from the assessment are reported in Tables 6.7-6.10 and in Figures 6.7-6.11.

The comparison of runs with and without tuning indices indicates is shown in Figure 6.12b. The information contains in both indices are consistent.

In 2022, the assessment shows a downward revision in F (Figure 6.12a). The estimate of fishing mortality is highly sensitive to the additional annual data, this is due to the low stock size.

Mohn's rho analysis (i.e. a measure of the relative difference between an estimate from an assessment with a truncated time-series and an estimate of the same quantity from an assessment using the full time-series) resulted in values of $-12 \%$ for $\mathrm{Fbar}_{(2-5)}, 23 \%$ for SSB and $40 \%$ for recruitment.

The retrospective bias in assessment when an additional year of data are incorporated may be due to the variability of cod recruitment over years, the strong dependency of the fishery to recruitment (not well estimated by the survey) and the unexpected disappearance of fish of older age.

Despite the high values of the Mohn's rho coefficient and the uncertainties in the estimates of the most recent year, the assessment has been validated (the stock is maintained in category 1 ), and the output are used to provide the short-term forecast. This decision follows the guidelines provide by WGBIAS (decision tree). Despite the uncertainties in the estimates of the most recent years, SSB and F are estimated well below biological references points.
The conclusions of the very recent benchmark was that given the recruitment driven dynamics of the stock and the low stock size reducing.

## State of the stock

Tables 6.7 and 6.8 summarise the estimated fishing mortality-at-age and the stock numbers-atage, respectively. The stock summary is reported in Table 6.9 and Figure 6.11.

Catches were around 5000 t between 2000 and 2016, with some higher catches following strong recruitments, and decreased around 1300 t since 2019 (Figure 6.11). Reliable discard estimates are available since 2004 and range between 134 and 3749 t depending on the interplay between recruitment dynamics and TAC constraints.

Recruitment has been highly variable over time with occasional very high recruitment followed by period of low recruitments. Since 2012, recruitment has been very weak with the exception of the 2014 year class, which is above average (Table 6.9 and Figure 6.11).

Spawning-stock biomass (SSB) has been fluctuating around $B_{\text {pa }}$ since 2004, except from 2011 to 2013 (as the consequence of a very good recruitment year) and is below Blim since 2017 (Table 6.9 and Figure 6.11, ICES, 2012).

Fishing mortality has been above $\mathrm{F}_{\text {mSY }}$ for the entire time-series, fluctuating between $\mathrm{F}_{\text {lim }}$ and $\mathrm{F}_{\mathrm{pa}}$. Fishing mortality increased up to above Flim between 2017 and 2019 (Table 6.9 and Figure 6.11).

### 6.5 Short-term projections

Assumptions made for the short-term projections are described in Table 6.12 and followed the stock annex.

F status quo was used as an assumption of $F$ in 2021 to reflect recent fishing pressure and was kept for 2022.

Recruitment values of 2022 and 2023 are similar in the stochastic forecast, because random resampling of a distribution may lead to identical median estimates. The recruitment age 1 fish values are 1305 thousands in 2022 and 2023.

SSB is predicted to be 992 t in 2023 which would still be below $\mathrm{Blim}_{\lim }(4200 \mathrm{t})$ (Table 6.11).
ICES provides zero-catch advice for this stock in 2023, because the median SSB remains below Blim by 2023 under all catch scenarios (Tables 6.12 and 6.13).

In the ICES advice framework, this would result in advised catches between 46 tonnes (at FMSY lower $\times$ SSB $_{2023} / \mathrm{MSY} \mathrm{B}_{\text {trigger }}$ ) and 77 tonnes (at $\mathrm{F}_{\text {MSY }} \times \mathrm{SSB}_{2023} / \mathrm{MSY} \mathrm{B}_{\text {trigger }}$ ), but the median SSB would remain below $\mathrm{B}_{\lim }$ by 2024.

The assumed recruitment in 2022 and 2023 used in the forecast constitutes a significant part (70\%) of the projected SSB in 2024 ( $45 \%$ and $25 \%$, respectively; Figure 6.14 and Table 6.14).

### 6.6 Medium-term projection

No medium-term projections were carried out.

### 6.7 Biological reference points

The reference points have been estimated using the agreed ICES guidelines, see Table 6.11 (ICES, 2016). $\mathrm{F}_{\mathrm{pa}}$ was set to $\mathrm{F}_{\mathrm{p} 0.5}$; the F that leads to $\mathrm{SSB} \geq \mathrm{Blim}_{\lim }$ with $95 \%$ probability at the last benchmark in 2020.

### 6.8 Management plans

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to demersal stocks including cod in ICES divisions 7.e-k.

### 6.9 Uncertainties and bias in assessment and forecast

The stock was benchmarked in 2020 (ICES, 2020a). The model was changed to a stochastic state -space assessment model (SAM). Maturity and natural mortality information was updated, discards were included in the assessment, catch (landings and discards) time-series were reviewed and updated from 2004 to 2018, commercial tuning series were reviewed and included as
biomass index, and survey indices were updated to a single modelled time-series using a vectorautoregressive spatio-temporal model (VAST). The F-pattern shows less variability across the time-series and higher estimates in most recent years than the previous assessment. Fishing mortality is observed to be sensitive to the addition of an extra year of data.

However, despite this uncertainty, it is quite clear that the cod stock is well below SBB limits and well above F target. Given that situation and the recommendations of WKBIAS, the last benchmark and WGCSE 2022 validated the proposed assessment model and its use for prediction.

### 6.10 Recommendations for future developments

There is room for development of a modelled commercial tuning fleet instead of the current method based on catch thresholds. Indeed, despite the work performed to improve the commercial tuning fleet, it is never easy to account for changes in fisheries targeting behaviours. Indeed, in recent years, cod is not targeted anymore by most of the fisheries.
Even if the survey index combined two surveys, it is based on few fish. Further work and sensitivity analysis on the VAST assumptions might also be performed and documented in the future to ensure that the model will converge for all ages and show low retrospective patterns.

### 6.11 Management considerations

The strong retrospective pattern implies that the current F estimates might be uncertain. Forecasts are sensitive to the assumption on recruitment as the landings are usually composed of a high proportion of age 2 fish (and age 1 for discards).

The recent technical measures introduced in the Celtic Sea, increase in the mesh size of the square mesh panels and raised lines are expected to reduce catches of Celtic Sea cod and improved the selection pattern. Impact of this measure should be monitored.

Additionally, mixed-fisheries issues could be responsible for maintaining $F$ at high level, as other gadoids fishing opportunities are higher. In this context, cod is no longer a target species but can be considered as bycatch in the fleet targeting haddock, whiting and Nephrops.

Historical information on management consideration can be found in the stock annex.

### 6.12 References

ICES. 2012. Report of the Working Group on the Celtic Seas Ecoregion (WGCSE), 9-18 May 2012, Copenhagen, Denmark. ICES CM 2012/ACOM:12.

ICES. 2016. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

Table 6.1. Cod in Division 7.e-k. History of official commercial landings presented by country and used by the Working Group. All weights are in tonnes.

| Year | Belgium | France | Ireland | UK | Others | Total | Discard estimates | Landings taken or reported in rectangles 33E2 and 33E3 * |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | NA | NA | NA | NA | NA | 5782 | NA | NA |
| 1972 | NA | NA | NA | NA | NA | 4737 | NA | NA |
| 1973 | NA | NA | NA | NA | NA | 4015 | NA | NA |
| 1974 | NA | NA | NA | NA | NA | 2898 | NA | NA |
| 1975 | NA | NA | NA | NA | NA | 3993 | NA | NA |
| 1976 | NA | NA | NA | NA | NA | 4818 | NA | NA |
| 1977 | NA | NA | NA | NA | NA | 3059 | NA | NA |
| 1978 | NA | NA | NA | NA | NA | 3647 | NA | NA |
| 1979 | NA | NA | NA | NA | NA | 4650 | NA | NA |
| 1980 | NA | NA | NA | NA | NA | 7243 | NA | NA |
| 1981 | NA | NA | NA | NA | NA | 10597 | NA | NA |
| 1982 | NA | NA | NA | NA | NA | 8766 | NA | NA |
| 1983 | NA | NA | NA | NA | NA | 9641 | NA | NA |
| 1984 | NA | NA | NA | NA | NA | 6631 | NA | NA |
| 1985 | NA | NA | NA | NA | NA | 8317 | NA | NA |
| 1986 | NA | NA | NA | NA | NA | 10475 | NA | NA |


| Year | Belgium | France | Ireland | UK | Others | Total | Discard estimates | Landings taken or reported in rectangles 33E2 and 33E3 * |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | NA | NA | NA | NA | NA | 10228 | NA | NA |
| 1988 | 554 | 13863 | 1480 | 1292 | 2 | 17191 | NA | NA |
| 1989 | 910 | 15801 | 1860 | 1223 | 15 | 19809 | NA | NA |
| 1990 | 621 | 9383 | 1241 | 1346 | 158 | 12749 | NA | NA |
| 1991 | 303 | 6260 | 1659 | 1094 | 20 | 9336 | NA | NA |
| 1992 | 195 | 7120 | 1212 | 1207 | 13 | 9747 | NA | NA |
| 1993 | 391 | 8317 | 766 | 945 | 6 | 10425 | NA | NA |
| 1994 | 398 | 7692 | 1616 | 906 | 8 | 10620 | NA | NA |
| 1995 | 400 | 8321 | 1946 | 1034 | 8 | 11709 | NA | NA |
| 1996 | 552 | 8981 | 1982 | 1166 | 0 | 12681 | NA | NA |
| 1997 | 694 | 8662 | 1513 | 1166 | 0 | 12035 | NA | NA |
| 1998 | 528 | 8096 | 1718 | 1089 | 0 | 11431 | NA | NA |
| 1999 | 326 | 5488 | 1883 | 897 | 0 | 8594 | NA | NA |
| 2000 | 208 | 4281 | 1302 | 744 | 0 | 6535 | NA | NA |
| 2001 | 347 | 6033 | 1091 | 838 | 0 | 8309 | NA | NA |
| 2002 | 555 | 7368 | 694 | 618 | 0 | 9235 | NA | NA |
| 2003 | 136 | 5222 | 517 | 346 | 0 | 6221 | NA | NA |


| Year | Belgium | France | Ireland | UK | Others | Total | Discard estimates | Landings taken or reported in rectangles 33E2 and 33E3 * |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 153 | 2934 | 657 | 281 | 1 | 4027 | 543 | 108 |
| 2005 | 186 | 2127 | 855 | 309 | 1 | 3478 | 1426 | 54 |
| 2006 | 101 | 2431 | 995 | 371 | 3 | 3902 | 2118 | 103 |
| 2007 | 107 | 3113 | 1208 | 411 | 3 | 4842 | 1248 | 527 |
| 2008 | 65 | 2994 | 1222 | 295 | 1 | 4577 | 306 | 558 |
| 2009 | 48 | 3020 | 847 | 267 | 5 | 4187 | 1229 | 193 |
| 2010 | 52 | 2449 | 1030 | 296 | 3 | 3831 | 3040 | 143 |
| 2011 | 123 | 4808 | 1010 | 427 | 7 | 6376 | 3749 | 147 |
| 2012 | 290 | 6900 | 1539 | 706 | 8 | 9443 | 2341 | 85 |
| 2013 | 202 | 5051 | 1470 | 548 | 3 | 7273 | 562 | 76 |
| 2014 | 141 | 2715 | 1189 | 466 | 0 | 4512 | 1569 | 24 |
| 2015 | 121 | 3373 | 1109 | 422 | 3 | 5028 | 483 | 39 |
| 2016 | 97 | 2579 | 881 | 365 | 1 | 3924 | 525 | 40 |
| 2017 | 82 | 1578 | 623 | 188 | 0 | 2471 | 134 | 19 |
| 2018 | 49 | 611 | 706 c | 130 | 0 | 1496 c | 316 | 20 |
| 2019 | 43 | 369 | 554 c | 84 | NA | 1051 c | 300 | 37 |
| 2020** | 18 | 371 | 487 | 44 | 2 | 922 | 231 | 71 |


| Year | Belgium | France | Ireland | UK | Others | Total | Discard estimates | Landings taken or reported in rectangles 33E2 and 33E3 * |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $2021^{* *}$ | 11 | 261 | 309 | 46 | 0 | 627 | 733 | 52 |

*Included in Ireland landings estimates. Landings in the south of Division 7.a (33E2 and 33E3) are included in the assessment and are considered to be part of the stock.
** Preliminary official landings.
c Incomplete due to part of the data being unavailable under national GDPR clauses.

Table 6.2. Cod in Division 7e-k. Catch number-at-age (in thousands). Number-at-age 1 and 2 before 2004 are estimated by the assessment model.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | NA | NA | 285 | 175 | 52 | 55 | 14 |
| 1981 | NA | NA | 811 | 153 | 41 | 20 | 12 |
| 1982 | NA | NA | 888 | 169 | 36 | 19 | 5 |
| 1983 | NA | NA | 540 | 424 | 77 | 21 | 11 |
| 1984 | NA | NA | 134 | 97 | 94 | 22 | 5 |
| 1985 | NA | NA | 465 | 61 | 40 | 47 | 15 |
| 1986 | NA | NA | 673 | 254 | 30 | 31 | 17 |
| 1987 | NA | NA | 448 | 250 | 62 | 20 | 15 |
| 1988 | NA | NA | 320 | 133 | 46 | 21 | 8 |
| 1989 | NA | NA | 2483 | 149 | 77 | 18 | 11 |
| 1990 | NA | NA | 1006 | 663 | 79 | 21 | 16 |
| 1991 | NA | NA | 229 | 330 | 203 | 48 | 14 |
| 1992 | NA | NA | 329 | 64 | 70 | 53 | 17 |
| 1993 | NA | NA | 928 | 79 | 24 | 19 | 16 |
| 1994 | NA | NA | 1199 | 258 | 27 | 10 | 17 |
| 1995 | NA | NA | 310 | 284 | 73 | 13 | 5 |
| 1996 | NA | NA | 1199 | 134 | 95 | 43 | 4 |
| 1997 | NA | NA | 951 | 297 | 48 | 22 | 6 |
| 1998 | NA | NA | 641 | 254 | 99 | 36 | 8 |
| 1999 | NA | NA | 756 | 158 | 59 | 36 | 14 |
| 2000 | NA | NA | 419 | 169 | 44 | 17 | 14 |
| 2001 | NA | NA | 136 | 98 | 70 | 19 | 19 |
| 2002 | NA | NA | 883 | 64 | 33 | 12 | 11 |
| 2003 | NA | NA | 827 | 217 | 15 | 9 | 7 |
| 2004 | 873 | 1077 | 229 | 189 | 65 | 5 | 6 |
| 2005 | 2875 | 2080 | 182 | 93 | 47 | 19 | 8 |
| 2006 | 7477 | 1052 | 295 | 17 | 25 | 13 | 9 |
| 2007 | 3556 | 1302 | 355 | 79 | 10 | 8 | 11 |


| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 467 | 885 | 403 | 122 | 27 | 4 | 6 |
| 2009 | 2212 | 421 | 424 | 120 | 47 | 11 | 4 |
| 2010 | 9794 | 618 | 151 | 107 | 46 | 14 | 5 |
| 2011 | 2325 | 4905 | 423 | 49 | 34 | 13 | 4 |
| 2012 | 746 | 1860 | 1757 | 117 | 18 | 14 | 11 |
| 2013 | 388 | 383 | 581 | 516 | 55 | 16 | 7 |
| 2014 | 4708 | 415 | 83 | 132 | 149 | 8 | 2 |
| 2015 | 242 | 2272 | 137 | 26 | 47 | 37 | 7 |
| 2016 | 624 | 195 | 707 | 33 | 7 | 17 | 16 |
| 2017 | 159 | 561 | 57 | 166 | 24 | 5 | 15 |
| 2018 | 902 | 172 | 137 | 14 | 38 | 5 | 2 |
| 2019 | 944 | 247 | 29 | 26 | 4 | 11 | 2 |
| 2020 | 342 | 548 | 36 | 3 | 2 | 2 | 2 |
| 2021 | 329 | 321 | 140 | 16 | 4 | 2 | 1 |

Table 6.3. Cod in Division 7e-k. Catch weight-(in kg) at-age.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.73800 |
| 1981 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.82433 |
| 1982 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.84160 |
| 1983 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 13.04373 |
| 1984 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.94520 |
| 1985 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.85860 |
| 1986 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.73800 |
| 1987 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.87613 |
| 1988 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 13.06075 |
| 1989 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.90245 |
| 1990 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 13.02887 |
| 1991 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.84900 |
| 1992 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.76847 |
| 1993 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.80275 |
| 1994 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.92082 |
| 1995 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 13.04880 |
| 1996 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.86750 |
| 1997 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.73800 |
| 1998 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.86750 |
| 1999 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.92300 |
| 2000 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.81200 |
| 2001 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.94226 |
| 2002 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.99664 |
| 2003 | 0.457 | 1.756 | 4.217 | 7.147 | 9.454 | 11.179 | 12.81200 |
| 2004 | 0.585 | 0.939 | 4.268 | 6.849 | 9.207 | 12.192 | 11.86933 |
| 2005 | 0.388 | 0.899 | 3.412 | 6.107 | 9.138 | 11.017 | 11.43300 |
| 2006 | 0.285 | 1.780 | 4.758 | 6.971 | 9.341 | 11.119 | 12.42300 |
| 2007 | 0.362 | 1.738 | 4.412 | 7.943 | 9.953 | 12.043 | 13.20200 |
| 2008 | 0.541 | 1.925 | 4.105 | 7.337 | 9.483 | 11.220 | 12.64783 |


| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 0.510 | 2.457 | 4.324 | 6.740 | 9.252 | 10.707 | 12.93800 |
| 2010 | 0.330 | 2.078 | 5.223 | 7.863 | 10.056 | 12.290 | 13.78180 |
| 2011 | 0.358 | 1.381 | 3.740 | 7.774 | 10.314 | 11.531 | 13.02500 |
| 2012 | 0.488 | 1.532 | 4.108 | 7.276 | 10.386 | 12.096 | 13.87391 |
| 2013 | 0.655 | 2.471 | 4.019 | 6.976 | 8.088 | 9.991 | 12.55800 |
| 2014 | 0.448 | 2.281 | 4.988 | 7.353 | 10.180 | 11.432 | 14.80600 |
| 2015 | 0.367 | 1.608 | 4.230 | 7.952 | 10.087 | 11.147 | 12.53600 |
| 2016 | 0.706 | 1.787 | 4.175 | 7.386 | 9.619 | 11.556 | 12.35400 |
| 2017 | 0.393 | 1.532 | 3.414 | 6.517 | 7.630 | 9.563 | 11.09620 |
| 2018 | 0.444 | 1.927 | 4.076 | 6.160 | 9.081 | 9.780 | 13.23200 |
| 2019 | 0.465 | 1.774 | 4.203 | 7.223 | 9.815 | 10.576 | 11.95100 |
| 2020 | 0.455 | 1.369 | 4.233 | 8.058 | 9.731 | 12.757 | 13.13100 |
| 2021 | 0.450 | 1.477 | 3.946 | 6.784 | 9.264 | 11.004 | 12.535 |

Table 6.4. Cod in Division 7e-k. Stock weight at age =1st quarter values.

| year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.574000 |
| 1981 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.578000 |
| 1982 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.578800 |
| 1983 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.626820 |
| 1984 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.583600 |
| 1985 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.607930 |
| 1986 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.574000 |
| 1987 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.580400 |
| 1988 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.695250 |
| 1989 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.620270 |
| 1990 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.640620 |
| 1991 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.579140 |
| 1992 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.575410 |
| 1993 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.577000 |
| 1994 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.582470 |
| 1995 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.588400 |
| 1996 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.580000 |
| 1997 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.574000 |
| 1998 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.580000 |
| 1999 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.582570 |
| 2000 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.577430 |
| 2001 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.605840 |
| 2002 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.624640 |
| 2003 | 0.370 | 1.421 | 3.936 | 6.901 | 9.324 | 11.107 | 13.577430 |
| 2004 | 0.356 | 0.830 | 4.035 | 6.101 | 9.324 | 13.784 | 9.952167 |
| 2005 | 0.320 | 0.830 | 4.035 | 6.101 | 9.324 | 11.135 | 15.169000 |
| 2006 | 0.267 | 1.516 | 4.370 | 6.325 | 9.350 | 11.081 | 12.688000 |
| 2007 | 0.290 | 1.453 | 3.916 | 8.101 | 10.658 | 11.413 | 15.827000 |
| 2008 | 0.344 | 1.623 | 4.027 | 7.200 | 8.941 | 10.916 | 12.550670 |


| year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 0.399 | 1.914 | 3.880 | 6.404 | 8.898 | 10.507 | 13.964000 |
| 2010 | 0.286 | 1.597 | 4.874 | 7.466 | 9.852 | 11.254 | 13.545200 |
| 2011 | 0.324 | 1.030 | 3.478 | 8.051 | 10.251 | 11.355 | 15.493000 |
| 2012 | 0.410 | 1.289 | 3.641 | 6.979 | 9.704 | 12.111 | 15.844000 |
| 2013 | 0.440 | 1.774 | 3.746 | 6.854 | 7.334 | 9.330 | 12.844000 |
| 2014 | 0.363 | 1.762 | 4.109 | 6.762 | 10.082 | 11.634 | 15.360000 |
| 2015 | 0.428 | 1.202 | 4.326 | 8.210 | 10.337 | 11.508 | 14.311000 |
| 2016 | 0.618 | 1.542 | 3.622 | 7.110 | 10.048 | 11.707 | 13.416000 |
| 2017 | 0.335 | 1.337 | 3.313 | 6.189 | 7.249 | 9.651 | 10.962330 |
| 2018 | 0.376 | 1.617 | 3.675 | 5.655 | 8.508 | 9.223 | 12.240000 |
| 2019 | 0.366 | 1.509 | 3.821 | 7.254 | 9.725 | 10.795 | 11.486000 |
| 2020 | 0.420 | 1.200 | 3.705 | 8.174 | 10.286 | 13.407 | 13.634000 |
| 2021 | 0.401 | 1.154 | 3.272 | 6.038 | 8.786 | 11.148 | 15.225 |

Table 6.5a. Cod in Division 7e-k. LPUE for French OT-DEF fleets. Units: landings in tonnes, effort in 000s hours fished and LPUE in $\mathrm{kg} /$ hour fished. This series is used to tuned the assessment model.

| Year | Effort | Landings |
| :--- | :---: | :---: |
| 2002 | 264146 | 3692073 |
| 2003 | 240535 | 1978251 |
| 2004 | 214247 | 918840 |
| 2005 | 156961 | 714850 |
| 2006 | 125245 | 712566 |
| 2007 | 150288 | 1193033 |
| 2008 | 138626 | 143812 |

Table 6.5b. Cod in Division 7e-k. Time-series of landings, effort and LPUE for the Irish fleets. Units: landings in tonnes live weight, effort in 000s hours fished and LPUE in kg/hour fished.

|  | Otter_trawl_27.7j |  |  | Beam_trawl_27.7j |  |  | Scottish_seiner_27.7j |  |  | Gillnet_27.7j |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landings | Effort | Ipue | Landings | Effort | Ipue | Landings | Effort | Ipue | Landings | Effort | Ipue |
| 1995 | 339,3 | 93,2 | 3,6 | 0,0 | 0,2 | 0,2 | 75,5 | 5,3 | 14,4 | 178,8 | 21,3 | 8,4 |
| 1996 | 326,4 | 70,2 | 4,6 | 8,7 | 1,4 | 6,3 | 124,5 | 8,2 | 15,3 | 65,0 | 5,2 | 12,4 |
| 1997 | 352,7 | 82,7 | 4,3 | 3,4 | 1,7 | 2,0 | 115,8 | 10,7 | 10,8 | 45,5 | 8,3 | 5,5 |
| 1998 | 262,7 | 89,1 | 2,9 | 19,1 | 5,2 | 3,7 | 103,4 | 6,6 | 15,6 | 59,1 | 16,0 | 3,7 |
| 1999 | 76,7 | 40,5 | 1,9 | 27,5 | 7,4 | 3,7 | 9,6 | 1,4 | 6,8 | 24,6 | 8,7 | 2,8 |
| 2000 | 95,5 | 63,9 | 1,5 | 21,2 | 6,9 | 3,1 | 24,4 | 3,5 | 7,0 | 13,8 | 7,0 | 2,0 |
| 2001 | 148,5 | 67,4 | 2,2 | 10,7 | 3,0 | 3,6 | 31,3 | 4,4 | 7,1 | 14,8 | 6,6 | 2,3 |
| 2002 | 150,0 | 90,4 | 1,7 | 5,4 | 3,1 | 1,7 | 24,6 | 8,9 | 2,8 | 12,3 | 8,1 | 1,5 |
| 2003 | 73,6 | 107,4 | 0,7 | 8,8 | 9,0 | 1,0 | 12,0 | 7,9 | 1,5 | 6,3 | 11,2 | 0,6 |
| 2004 | 36,1 | 88,3 | 0,4 | 2,5 | 2,2 | 1,2 | 10,3 | 8,1 | 1,3 | 4,2 | 6,1 | 0,7 |
| 2005 | 37,8 | 71,3 | 0,5 | 4,7 | 2,4 | 2,0 | 17,5 | 5,8 | 3,0 | 3,4 | 6,1 | 0,6 |
| 2006 | 39,6 | 64,5 | 0,6 | 2,0 | 1,5 | 1,3 | 15,6 | 5,3 | 2,9 | 7,2 | 7,3 | 1,0 |
| 2007 | 35,9 | 78,3 | 0,5 | 7,8 | 2,4 | 3,3 | 9,8 | 3,5 | 2,8 | 6,5 | 10,5 | 0,6 |
| 2008 | 33,1 | 66,7 | 0,5 | 2,6 | 1,1 | 2,3 | 9,5 | 2,8 | 3,3 | 6,5 | 7,9 | 0,8 |
| 2009 | 26,6 | 73,0 | 0,4 | 4,7 | 2,8 | 1,7 | 8,9 | 3,3 | 2,7 | 8,0 | 10,9 | 0,7 |
| 2010 | 52,5 | 85,7 | 0,6 | 1,7 | 1,0 | 1,7 | 17,0 | 4,4 | 3,9 | 8,4 | 9,4 | 0,9 |
| 2011 | 57,7 | 62,8 | 0,9 | 1,7 | 0,6 | 2,7 | 21,6 | 4,6 | 4,7 | 16,8 | 8,0 | 2,1 |
| 2012 | 62,8 | 65,6 | 1,0 | 0,4 | 0,3 | 1,5 | 29,8 | 5,4 | 5,6 | 25,2 | 8,3 | 3,0 |
| 2013 | 66,1 | 61,3 | 1,1 | 1,8 | 0,6 | 3,3 | 32,5 | 6,6 | 4,9 | 15,4 | 9,8 | 1,6 |
| 2014 | 51,6 | 53,9 | 1,0 | 1,2 | 0,6 | 1,9 | 52,6 | 7,4 | 7,1 | 9,7 | 12,2 | 0,8 |
| 2015 | 63,6 | 46,9 | 1,4 | 0,6 | 0,1 | 6,3 | 38,2 | 5,3 | 7,2 | 18,1 | 14,2 | 1,3 |
| 2016 | 48,5 | 50,7 | 1,0 | 0,3 | 0,2 | 1,5 | 25,2 | 5,3 | 4,7 | 15,8 | 17,1 | 0,9 |
| 2017 | 41,3 | 56,4 | 0,7 | 0,0 | 0,0 | 10,0 | 24,0 | 5,3 | 4,5 | 10,4 | 18,0 | 0,6 |
| 2018 | 42,3 | 52,1 | 0,8 | 0,2 | 0,1 | 2,4 | 28,5 | 6,4 | 4,5 | 5,9 | 16,8 | 0,4 |
| 2019 | 30,4 | 53,4 | 0,6 | 0,2 | 0,1 | 1,7 | 18,1 | 7,3 | 2,5 | 5,5 | 14,5 | 0,4 |
| 2020 | 26,72 | 44,11 | 0,61 | 0,07 | 0,02 | 2,92 | 17,16 | 5,53 | 3,10 | 13,80 | 13,52 | 1,02 |
| 2021 | 12,95 | 40,28 | 0,32 | 0,11 | 0,31 | 0,35 | 11,05 | 5,29 | 2,09 | 15,24 | 19,73 | 0,77 |


| NA | Otter_trawl_27.7g |  |  | Beam_trawl_27.7g |  |  | Scottish_seiner_27.7g |  |  | Gilnet_27.7g |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landings | Effort | Ipue | Landings | Effort | Ipue | Landings | Effort | Ipue | Landings | Effort | Ipue |
| 1995 | 429,8 | 63,3 | 6,8 | 85,8 | 20,7 | 4,1 | 111,3 | 6,4 | 17,3 | 114,9 | 6,3 | 18,1 |
| 1996 | 569,2 | 60,0 | 9,5 | 112,5 | 26,7 | 4,2 | 164,9 | 9,7 | 16,9 | 338,9 | 6,2 | 54,8 |
| 1997 | 401,9 | 65,0 | 6,2 | 131,5 | 28,1 | 4,7 | 215,2 | 16,1 | 13,4 | 52,8 | 1,9 | 27,7 |
| 1998 | 450,5 | 72,3 | 6,2 | 166,8 | 35,2 | 4,7 | 264,1 | 14,9 | 17,7 | 87,3 | 3,4 | 25,4 |
| 1999 | 300,7 | 51,5 | 5,8 | 190,6 | 40,8 | 4,7 | 64,6 | 8,0 | 8,1 | 200,4 | 8,4 | 23,9 |
| 2000 | 279,4 | 60,6 | 4,6 | 180,6 | 36,8 | 4,9 | 106,0 | 9,9 | 10,8 | 151,7 | 10,1 | 15,0 |
| 2001 | 358,5 | 69,4 | 5,2 | 101,2 | 39,5 | 2,6 | 115,0 | 16,3 | 7,0 | 115,8 | 8,8 | 13,2 |
| 2002 | 212,9 | 77,2 | 2,8 | 57,9 | 31,5 | 1,8 | 71,0 | 20,9 | 3,4 | 31,0 | 6,4 | 4,8 |
| 2003 | 167,2 | 86,8 | 1,9 | 56,8 | 49,2 | 1,2 | 35,6 | 20,1 | 1,8 | 31,3 | 11,1 | 2,8 |
| 2004 | 190,2 | 97,1 | 2,0 | 74,3 | 54,9 | 1,4 | 54,4 | 18,4 | 3,0 | 62,0 | 13,5 | 4,6 |
| 2005 | 292,5 | 124,7 | 2,3 | 118,9 | 49,6 | 2,4 | 64,4 | 14,6 | 4,4 | 77,9 | 10,9 | 7,2 |
| 2006 | 379,4 | 118,0 | 3,2 | 128,6 | 60,5 | 2,1 | 91,0 | 14,8 | 6,2 | 63,7 | 7,8 | 8,1 |
| 2007 | 316,1 | 135,4 | 2,3 | 96,2 | 55,8 | 1,7 | 58,5 | 15,8 | 3,7 | 85,4 | 9,4 | 9,1 |
| 2008 | 344,9 | 125,4 | 2,7 | 85,4 | 37,2 | 2,3 | 55,6 | 11,6 | 4,8 | 88,0 | 14,1 | 6,2 |
| 2009 | 405,9 | 137,1 | 3,0 | 74,4 | 37,9 | 2,0 | 34,6 | 8,2 | 4,2 | 81,1 | 13,8 | 5,9 |
| 2010 | 524,8 | 140,8 | 3,7 | 94,7 | 40,2 | 2,4 | 54,3 | 9,7 | 5,6 | 76,0 | 14,0 | 5,4 |
| 2011 | 438,4 | 120,3 | 3,6 | 82,5 | 35,3 | 2,3 | 46,7 | 11,0 | 4,2 | 76,6 | 11,3 | 6,7 |
| 2012 | 780,7 | 127,7 | 6,1 | 161,9 | 40,3 | 4,0 | 111,5 | 14,1 | 7,9 | 129,1 | 15,4 | 8,4 |
| 2013 | 721,4 | 118,2 | 6,1 | 195,8 | 38,5 | 5,1 | 111,3 | 13,2 | 8,5 | 92,5 | 14,4 | 6,4 |
| 2014 | 600,1 | 127,3 | 4,7 | 142,9 | 37,8 | 3,8 | 110,5 | 12,5 | 8,9 | 59,2 | 14,1 | 4,2 |
| 2015 | 526,3 | 132,7 | 4,0 | 160,1 | 37,8 | 4,2 | 59,2 | 9,3 | 6,4 | 48,7 | 12,5 | 3,9 |
| 2016 | 418,1 | 148,2 | 2,8 | 106,8 | 39,6 | 2,7 | 51,1 | 10,4 | 4,9 | 47,1 | 13,6 | 3,5 |
| 2017 | 361,4 | 136,1 | 2,7 | 46,4 | 35,2 | 1,3 | 42,1 | 9,7 | 4,3 | 22,4 | 14,8 | 1,5 |
| 2018 | 387,6 | 108,2 | 3,6 | 72,6 | 37,4 | 1,9 | 61,1 | 9,7 | 6,3 | 16,7 | 14,0 | 1,2 |
| 2019 | 244,8 | 103,9 | 2,4 | 71,9 | 34,1 | 2,1 | 50,9 | 14,3 | 3,6 | 21,9 | 16,0 | 1,4 |
| 2020 | 184,36 | 89,91 | 2,05 | 55,00 | 29,14 | 1,89 | 51,51 | 13,59 | 3,79 | 20,08 | 15,02 | 1,34 |
| 2021 | 108,54 | 83,90 | 1,29 | 45,08 | 31,57 | 1,43 | 28,73 | 14,8 | 1,94 | 15,59 | 17,59 | 0,89 |

Table 6.5c. Cod in Division 7e-k. Time-series of landings, effort and LPUE for the UK fleets. Units: landings in tonnes, effort in days fished and LPUE in kg/day.

| YEAR | Beam_trawl_27.7ek |  | Trawl_27.7ek |  | Trawl_27.7e |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Lands..t. | Effort..Days. | Lands..t.. 1 | Effort..Days.. 1 | Lands..t.. 2 | Effort..Days.. 2 |
| 1983 | 25.55 | 2853 | 40.93 | 2573 | 20.60 | 1871 |
| 1984 | 128.75 | 8427 | 235.68 | 8092 | 76.42 | 5618 |
| 1985 | 145.39 | 7706 | 250.67 | 7186 | 63.97 | 5411 |
| 1986 | 165.76 | 6651 | 232.19 | 6174 | 78.31 | 4425 |
| 1987 | 248.91 | 8060 | 210.36 | 5446 | 88.49 | 3701 |
| 1988 | 249.21 | 9487 | 262.68 | 5645 | 151.35 | 4265 |
| 1989 | 231.24 | 10071 | 177.12 | 5997 | 96.00 | 4607 |
| 1990 | 309.07 | 10477 | 305.78 | 6661 | 119.41 | 4423 |
| 1991 | 256.19 | 9017 | 242.33 | 5938 | 83.60 | 4004 |
| 1992 | 256.33 | 8183 | 231.85 | 6494 | 80.76 | 4108 |
| 1993 | 221.79 | 9511 | 183.05 | 5055 | 42.88 | 3761 |
| 1994 | 179.13 | 13925 | 78.23 | 4426 | 41.25 | 3423 |
| 1995 | 241.35 | 15076 | 115.05 | 4405 | 55.09 | 3294 |
| 1996 | 304.22 | 15748 | 120.46 | 4476 | 59.21 | 2589 |
| 1997 | 303.67 | 16373 | 150.01 | 5088 | 79.81 | 3011 |
| 1998 | 266.15 | 15574 | 119.56 | 4729 | 62.50 | 2699 |
| 1999 | 257.43 | 15614 | 90.68 | 6638 | 46.81 | 2486 |
| 2000 | 188.07 | 16456 | 110.79 | 7054 | 52.59 | 2681 |
| 2001 | 257.24 | 17335 | 109.75 | 5875 | 59.05 | 2732 |
| 2002 | 132.13 | 16503 | 82.70 | 5657 | 34.11 | 2448 |
| 2003 | 108.77 | 18285 | 58.80 | 5120 | 24.48 | 2273 |
| 2004 | 96.93 | 18250 | 44.06 | 5273 | 15.05 | 2334 |
| 2005 | 103.60 | 17157 | 41.13 | 5047 | 17.38 | 1762 |
| 2006 | 91.88 | 15412 | 55.43 | 5314 | 13.54 | 1699 |
| 2007 | 111.28 | 15085 | 49.65 | 5679 | 21.61 | 1917 |
| 2008 | 71.38 | 13734 | 49.34 | 4686 | 24.26 | 1750 |
| 2009 | 67.27 | 12170 | 27.56 | 4928 | 12.56 | 1847 |
| 2010 | 65.62 | 12150 | 31.13 | 5185 | 15.27 | 2213 |


| YEAR | Beam_trawl_27.7ek |  | Trawl_27.7ek |  | Trawl_27.7e |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Lands..t. | Effort..Days. | Lands..t.. 1 | Effort..Days.. 1 | Lands..t.. 2 | Effort..Days.. 2 |
| 2011 | 99.03 | 13205 | 47.73 | 4354 | 26.00 | 1931 |
| 2012 | 165.63 | 13411 | 79.03 | 4312 | 30.95 | 2068 |
| 2013 | 114.49 | 12950 | 37.30 | 2014 | 22.94 | 1587 |
| 2014 | 87.55 | 12807 | 17.07 | 1606 | 14.06 | 1440 |
| 2015 | 89.39 | 12769 | 16.68 | 1061 | 14.40 | 978 |
| 2016 | 73.81 | 13913 | 0.00 | 0 | 0.00 | 0 |
| 2017 | 35.49 | 14283 | 19.37 | 3718 | 9.33 | 2398 |
| 2018 | 24.41 | 13065 | 17.51 | 3233 | 5.34 | 1987 |
| 2019 | 18.03 | 12649 | 11.76 | 2660 | 3.64 | 1548 |
| 2020 | 10.21 | 12332 | 2.55 | 1481 | 1.74 | 1093 |
| 2021 | 14.87 | 12593 | 2.53 | 1895 | 1.25 | 1353 |

Table 6.6. Cod in Division 7e-k. Time-series of survey indices scrutinized at WGCSE and used in the assessment.

| Cod in Divisions 7e-k, tuning fleets, WGCSE2021 |  |  |  |
| :---: | :---: | :---: | :---: |
| 102 |  |  |  |
| FR-OTDEF Q2+3+4 trawlers in 7e-k |  |  |  |
| 2002 | 2020 |  |  |
| 1 | 1 | 0.25 | 1 |
| -1 | -1 |  |  |
| Year | Effort | Landings |  |
| 2002 | 264146 | 3692073 |  |
| 2003 | 240535 | 1978251 |  |
| 2004 | 214247 | 918840 |  |
| 2005 | 156961 | 714850 |  |
| 2006 | 125245 | 712566 |  |
| 2007 | 150288 | 1193033 |  |
| 2008 | 138626 | 814340 |  |
| 2009 | 143812 | 647808 |  |
| 2010 | 143730 | 705691 |  |
| 2011 | 258383 | 2332986 |  |
| 2012 | 252110 | 3393990 |  |
| 2013 | 190886 | 1696287 |  |
| 2014 | 151518 | 1113363 |  |
| 2015 | 185791 | 1374691 |  |
| 2016 | 178399 | 1122665 |  |
| 2017 | 137849 | 483571 |  |
| 2018 | 102586 | 163178 |  |
| 2019 | 114838 | 136473 |  |
| 2020 | 96907 | 149412 |  |
| 2021 | 97502 | 102964 |  |
| next table |  |  |  |
| IR-GFS FR-EVHOE Q4 combined indices - VAST Modelling |  |  |  |
| 2003 | 2020 | NA |  |


| 1 |  |  | 0.79 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  |  | NA |  |  |
| Year | Effort | Age 1 | Age 2 | Age 3 | Age 4 |
| 2003 | 1 | 24.431 | 39.006 | 49.727 | 17.447 |
| 2004 | 1 | 34.942 | 33.287 | 15.157 | 18.076 |
| 2005 | 1 | 112.156 | 33.891 | 12.266 | 0.000 |
| 2006 | 1 | 74.788 | 44.300 | 8.044 | 0.000 |
| 2007 | 1 | 95.111 | 69.869 | 33.235 | 12.524 |
| 2008 | 1 | 29.186 | 72.709 | 30.874 | 11.022 |
| 2009 | 1 | 58.069 | 20.743 | 27.982 | 11.823 |
| 2010 | 1 | 491.426 | 62.255 | 5.542 | 7.331 |
| 2011 | 1 | 241.122 | 364.573 | 24.152 | 4.165 |
| 2012 | 1 | 21.254 | 115.420 | 141.296 | 26.223 |
| 2013 | 1 | 25.047 | 8.148 | 23.572 | 33.161 |
| 2014 | 1 | 292.211 | 30.564 | 13.048 | 20.528 |
| 2015 | 1 | 13.884 | 154.490 | 9.121 | 0.000 |
| 2016 | 1 | 128.255 | 21.528 | 113.529 | 17.994 |
| 2017 | 1 | 21.796 | 65.972 | 26.003 | 38.008 |
| 2018 | 1 | 36.502 | 9.271 | 12.465 | 11.512 |
| 2019 | 1 | 145.144 | 36.138 | 2.239 | 6.908 |
| 2020 | 1 | 55.313 | 105.361 | 3.107 | 0.958 |
| 2021 | 1 | 23.014 | 28.033 | 37.635 | 1.668 |

Table6.7. Cod in Division 7e-k. Final SAM fishing mortality-at-age.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ | $F_{\text {bar }}($ mean 2-5) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.494 | 0.976 | 0.912 | 0.891 | 0.880 | 1.107 | 1.107 | 0.915 |
| 1981 | 0.485 | 0.958 | 0.894 | 0.869 | 0.855 | 1.073 | 1.073 | 0.894 |
| 1982 | 0.461 | 0.909 | 0.845 | 0.818 | 0.802 | 1.002 | 1.002 | 0.843 |
| 1983 | 0.464 | 0.915 | 0.851 | 0.821 | 0.803 | 1.000 | 1.000 | 0.848 |
| 1984 | 0.426 | 0.836 | 0.775 | 0.743 | 0.727 | 0.902 | 0.902 | 0.770 |
| 1985 | 0.428 | 0.841 | 0.779 | 0.745 | 0.726 | 0.898 | 0.898 | 0.773 |
| 1986 | 0.450 | 0.887 | 0.824 | 0.786 | 0.763 | 0.936 | 0.936 | 0.815 |
| 1987 | 0.459 | 0.905 | 0.842 | 0.802 | 0.778 | 0.947 | 0.947 | 0.832 |
| 1988 | 0.442 | 0.869 | 0.808 | 0.764 | 0.741 | 0.897 | 0.897 | 0.796 |
| 1989 | 0.464 | 0.914 | 0.851 | 0.800 | 0.773 | 0.928 | 0.928 | 0.835 |
| 1990 | 0.500 | 0.990 | 0.924 | 0.867 | 0.835 | 0.995 | 0.995 | 0.904 |
| 1991 | 0.531 | 1.054 | 0.989 | 0.927 | 0.894 | 1.060 | 1.060 | 0.966 |
| 1992 | 0.531 | 1.055 | 0.990 | 0.925 | 0.893 | 1.056 | 1.056 | 0.966 |
| 1993 | 0.525 | 1.042 | 0.977 | 0.908 | 0.876 | 1.032 | 1.032 | 0.951 |
| 1994 | 0.538 | 1.068 | 1.002 | 0.929 | 0.896 | 1.053 | 1.053 | 0.974 |
| 1995 | 0.526 | 1.044 | 0.979 | 0.904 | 0.872 | 1.022 | 1.022 | 0.950 |
| 1996 | 0.533 | 1.057 | 0.992 | 0.909 | 0.874 | 1.017 | 1.017 | 0.958 |
| 1997 | 0.525 | 1.041 | 0.976 | 0.887 | 0.845 | 0.974 | 0.974 | 0.937 |
| 1998 | 0.538 | 1.068 | 1.003 | 0.906 | 0.859 | 0.980 | 0.980 | 0.959 |
| 1999 | 0.546 | 1.085 | 1.019 | 0.916 | 0.865 | 0.978 | 0.978 | 0.972 |
| 2000 | 0.541 | 1.075 | 1.010 | 0.903 | 0.851 | 0.955 | 0.955 | 0.960 |
| 2001 | 0.550 | 1.093 | 1.028 | 0.919 | 0.868 | 0.968 | 0.968 | 0.977 |
| 2002 | 0.562 | 1.117 | 1.052 | 0.934 | 0.879 | 0.975 | 0.975 | 0.995 |
| 2003 | 0.551 | 1.093 | 1.030 | 0.910 | 0.856 | 0.947 | 0.947 | 0.972 |
| 2004 | 0.539 | 1.069 | 1.008 | 0.889 | 0.838 | 0.928 | 0.928 | 0.951 |
| 2005 | 0.554 | 1.097 | 1.038 | 0.915 | 0.867 | 0.962 | 0.962 | 0.979 |
| 2006 | 0.519 | 1.023 | 0.969 | 0.855 | 0.818 | 0.914 | 0.914 | 0.916 |
| 2007 | 0.507 | 1.000 | 0.952 | 0.844 | 0.814 | 0.917 | 0.917 | 0.903 |
| 2008 | 0.486 | 0.959 | 0.920 | 0.822 | 0.801 | 0.909 | 0.909 | 0.875 |


| 2009 | 0.481 | 0.951 | 0.917 | 0.824 | 0.811 | 0.928 | 0.928 | 0.875 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2010 | 0.465 | 0.921 | 0.891 | 0.804 | 0.798 | 0.921 | 0.921 | 0.853 |
| 2011 | 0.462 | 0.918 | 0.888 | 0.801 | 0.798 | 0.927 | 0.927 | 0.851 |
| 2012 | 0.489 | 0.977 | 0.947 | 0.855 | 0.856 | 1.001 | 1.001 | 0.908 |
| 2013 | 0.508 | 1.023 | 0.994 | 0.900 | 0.906 | 1.065 | 1.065 | 0.956 |
| 2014 | 0.483 | 0.973 | 0.942 | 0.852 | 0.858 | 1.011 | 1.011 | 0.906 |
| 2015 | 0.495 | 1.000 | 0.967 | 0.873 | 0.880 | 1.044 | 1.044 | 0.930 |
| 2016 | 0.518 | 1.052 | 1.017 | 0.916 | 0.924 | 1.101 | 1.101 | 0.977 |
| 2017 | 0.587 | 1.201 | 1.162 | 1.048 | 1.058 | 1.264 | 1.264 | 1.117 |
| 2018 | 0.612 | 1.254 | 1.212 | 1.087 | 1.094 | 1.307 | 1.307 | 1.162 |
| 2019 | 0.600 | 1.231 | 1.188 | 1.066 | 1.072 | 1.289 | 1.289 | 1.139 |
| 2020 | 0.541 | 1.105 | 1.062 | 0.950 | 0.961 | 1.162 | 1.162 | 1.020 |
| 0254 | 1.133 | 1.089 | 0.977 | 0.990 | 1.196 | 1.196 | 1.047 |  |

Table 6.8. Cod in Division 7e-k. Final SAM stock number-at-age.

| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 17720 | 4867 | 698 | 299 | 89 | 70 | 21 |
| 1981 | 8019 | 6664 | 1318 | 228 | 95 | 29 | 23 |
| 1982 | 3889 | 2988 | 1869 | 401 | 76 | 34 | 13 |
| 1983 | 8433 | 1440 | 864 | 636 | 141 | 28 | 15 |
| 1984 | 9296 | 3258 | 402 | 271 | 216 | 51 | 12 |
| 1985 | 8141 | 3697 | 1030 | 149 | 104 | 84 | 22 |
| 1986 | 9219 | 3198 | 1152 | 373 | 61 | 43 | 34 |
| 1987 | 27983 | 3487 | 942 | 378 | 130 | 25 | 25 |
| 1988 | 14070 | 11124 | 993 | 314 | 123 | 47 | 15 |
| 1989 | 4807 | 5524 | 3482 | 365 | 120 | 46 | 21 |
| 1990 | 6490 | 1779 | 1604 | 1106 | 145 | 44 | 23 |
| 1991 | 17005 | 2352 | 462 | 478 | 347 | 57 | 21 |
| 1992 | 19077 | 6167 | 581 | 132 | 139 | 106 | 23 |
| 1993 | 10876 | 6892 | 1571 | 163 | 42 | 44 | 35 |
| 1994 | 21331 | 3797 | 1773 | 457 | 53 | 14 | 23 |
| 1995 | 16618 | 7699 | 915 | 488 | 143 | 18 | 10 |
| 1996 | 12144 | 5958 | 1987 | 286 | 151 | 49 | 8 |
| 1997 | 12659 | 4287 | 1490 | 569 | 102 | 47 | 15 |
| 1998 | 7921 | 4586 | 1080 | 420 | 188 | 41 | 18 |
| 1999 | 4374 | 2809 | 1145 | 299 | 128 | 65 | 19 |
| 2000 | 16676 | 1474 | 684 | 311 | 94 | 43 | 26 |
| 2001 | 14810 | 6063 | 347 | 189 | 98 | 33 | 23 |
| 2002 | 5412 | 5275 | 1506 | 99 | 60 | 31 | 18 |
| 2003 | 3354 | 1850 | 1196 | 386 | 31 | 20 | 15 |
| 2004 | 4741 | 1211 | 443 | 309 | 119 | 11 | 11 |
| 2005 | 7966 | 1865 | 277 | 136 | 98 | 41 | 8 |
| 2006 | 8275 | 2736 | 469 | 64 | 44 | 32 | 16 |
| 2007 | 6203 | 2819 | 729 | 145 | 21 | 16 | 17 |
| 2008 | 2242 | 2181 | 719 | 214 | 50 | 8 | 11 |


| Year | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 5403 | 847 | 612 | 221 | 74 | 19 | 6 |
| 2010 | 24640 | 1998 | 250 | 188 | 79 | 25 | 8 |
| 2011 | 9323 | 9490 | 615 | 88 | 66 | 29 | 10 |
| 2012 | 1771 | 3552 | 2762 | 203 | 34 | 24 | 13 |
| 2013 | 2337 | 638 | 958 | 816 | 70 | 13 | 11 |
| 2014 | 11992 | 852 | 169 | 276 | 266 | 22 | 6 |
| 2015 | 1111 | 4661 | 244 | 50 | 93 | 87 | 9 |
| 2016 | 2320 | 407 | 1207 | 76 | 16 | 31 | 27 |
| 2017 | 666 | 821 | 110 | 302 | 26 | 6 | 16 |
| 2018 | 1472 | 222 | 173 | 27 | 74 | 7 | 5 |
| 2019 | 3502 | 459 | 44 | 36 | 8 | 19 | 3 |
| 2020 | 1305 | 1216 | 86 | 10 | 8 | 2 | 4 |
| 2021 | 923 | 455 | 270 | 23 | 4 | 3 | 2 |

Table 8.9. Cod in Divisions 7e-k. Final SAM summary table.

| Year | R(age 1) | Low | High | SSB | Low | High | $F_{\text {bar(2-5) }}$ | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 17720 | 8502 | 36932 | 10247 | 7843 | 13387 | 0.915 | 0.749 | 1.119 | 20177 | 13640 | 29845 |
| 1981 | 8019 | 3988 | 16127 | 13034 | 9933 | 17105 | 0.894 | 0.750 | 1.066 | 20721 | 14781 | 29049 |
| 1982 | 3889 | 1949 | 7760 | 13167 | 10494 | 16521 | 0.843 | 0.712 | 0.999 | 17074 | 13373 | 21799 |
| 1983 | 8433 | 4312 | 16495 | 10489 | 8677 | 12681 | 0.848 | 0.716 | 1.003 | 14789 | 11776 | 18573 |
| 1984 | 9296 | 4767 | 18131 | 8588 | 6982 | 10565 | 0.770 | 0.630 | 0.942 | 14269 | 10651 | 19114 |
| 1985 | 8141 | 4205 | 15762 | 9830 | 7912 | 12213 | 0.773 | 0.640 | 0.932 | 15542 | 11741 | 20574 |
| 1986 | 9219 | 4695 | 18101 | 10758 | 8700 | 13301 | 0.815 | 0.693 | 0.958 | 16576 | 12729 | 21587 |
| 1987 | 27983 | 14496 | 54019 | 10551 | 8582 | 12972 | 0.832 | 0.707 | 0.978 | 23444 | 16386 | 33542 |
| 1988 | 14070 | 7264 | 27253 | 16219 | 12026 | 21873 | 0.796 | 0.659 | 0.960 | 28970 | 20206 | 41534 |
| 1989 | 4807 | 2454 | 9419 | 21417 | 16488 | 27820 | 0.835 | 0.705 | 0.988 | 27766 | 21184 | 36393 |
| 1990 | 6490 | 3257 | 12932 | 17023 | 13673 | 21193 | 0.904 | 0.773 | 1.058 | 21029 | 16842 | 26257 |
| 1991 | 17005 | 8528 | 33908 | 10934 | 9097 | 13143 | 0.966 | 0.811 | 1.150 | 18890 | 13951 | 25578 |
| 1992 | 19077 | 9725 | 37422 | 10555 | 8075 | 13797 | 0.966 | 0.824 | 1.132 | 21805 | 14990 | 31719 |
| 1993 | 10876 | 5449 | 21710 | 13519 | 10277 | 17783 | 0.951 | 0.818 | 1.106 | 22481 | 16162 | 31269 |
| 1994 | 21331 | 10979 | 41447 | 13521 | 10730 | 17038 | 0.974 | 0.834 | 1.137 | 24384 | 17768 | 33463 |
| 1995 | 16618 | 8616 | 32050 | 14295 | 11018 | 18546 | 0.950 | 0.818 | 1.102 | 25728 | 18429 | 35916 |
| 1996 | 12144 | 6327 | 23307 | 15874 | 12485 | 20184 | 0.958 | 0.825 | 1.114 | 24810 | 18594 | 33104 |
| 1997 | 12659 | 6588 | 24326 | 14352 | 11592 | 17770 | 0.937 | 0.805 | 1.091 | 22249 | 16992 | 29133 |
| 1998 | 7921 | 4120 | 15232 | 12818 | 10363 | 15855 | 0.959 | 0.826 | 1.114 | 19044 | 14555 | 24918 |
| 1999 | 4374 | 2287 | 8365 | 10582 | 8628 | 12978 | 0.972 | 0.835 | 1.130 | 14352 | 11320 | 18197 |
| 2000 | 16676 | 9099 | 30560 | 7486 | 6228 | 8999 | 0.960 | 0.826 | 1.115 | 14808 | 10932 | 20059 |
| 2001 | 14810 | 8194 | 26768 | 8824 | 6725 | 11577 | 0.977 | 0.844 | 1.132 | 18362 | 13296 | 25358 |
| 2002 | 5412 | 3108 | 9426 | 11384 | 9147 | 14167 | 0.995 | 0.857 | 1.156 | 17249 | 13446 | 22129 |
| 2003 | 3354 | 2123 | 5296 | 9163 | 7727 | 10867 | 0.972 | 0.847 | 1.116 | 11943 | 10025 | 14227 |
| 2004 | 4741 | 2937 | 7653 | 5463 | 4694 | 6358 | 0.951 | 0.838 | 1.080 | 7738 | 6566 | 9119 |
| 2005 | 7966 | 5530 | 11477 | 4190 | 3640 | 4823 | 0.979 | 0.849 | 1.129 | 7530 | 6393 | 8868 |
| 2006 | 8275 | 5536 | 12371 | 5522 | 4708 | 6477 | 0.916 | 0.809 | 1.037 | 9783 | 8177 | 11705 |
| 2007 | 6203 | 4238 | 9079 | 6708 | 5695 | 7900 | 0.903 | 0.796 | 1.023 | 10591 | 8897 | 12607 |
| 2008 | 2242 | 1541 | 3262 | 6814 | 5818 | 7981 | 0.875 | 0.769 | 0.996 | 9416 | 8013 | 11065 |


| Year | R(age 1) | Low | High | SSB | Low | High | Fbar(2-5) | Low | High | TSB | Low | High |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2009 | 5403 | 3739 | 7808 | 5438 | 4656 | 6350 | 0.875 | 0.771 | 0.994 | 8505 | 7273 | 9947 |
| 2010 | 24640 | 17401 | 34892 | 5431 | 4675 | 6310 | 0.853 | 0.742 | 0.981 | 14031 | 11440 | 17211 |
| 2011 | 9323 | 6505 | 13361 | 9140 | 7692 | 10861 | 0.851 | 0.739 | 0.980 | 16807 | 13986 | 20196 |
| 2012 | 1771 | 1229 | 2551 | 14076 | 11764 | 16842 | 0.908 | 0.806 | 1.024 | 17612 | 14870 | 20859 |
| 2013 | 2337 | 1599 | 3414 | 10315 | 8747 | 12165 | 0.956 | 0.831 | 1.100 | 12115 | 10433 | 14068 |
| 2014 | 11992 | 8271 | 17387 | 6343 | 5420 | 7423 | 0.906 | 0.791 | 1.038 | 11435 | 9560 | 13678 |
| 2015 | 1111 | 759 | 1627 | 6499 | 5510 | 7666 | 0.930 | 0.813 | 1.064 | 9626 | 7946 | 11660 |
| 2016 | 2320 | 1586 | 3394 | 5834 | 4832 | 7043 | 0.977 | 0.856 | 1.117 | 7862 | 6625 | 9331 |
| 2017 | 666 | 447 | 992 | 3224 | 2705 | 3843 | 1.117 | 0.972 | 1.285 | 3978 | 3364 | 4703 |
| 2018 | 1472 | 1000 | 2168 | 1689 | 1450 | 1966 | 1.162 | 1.004 | 1.345 | 2452 | 2113 | 2845 |
| 2019 | 3502 | 2423 | 5062 | 1097 | 940 | 1279 | 1.139 | 0.984 | 1.320 | 2709 | 2202 | 3333 |
| 2020 | 1305 | 856 | 1990 | 1346 | 1105 | 1640 | 1.020 | 0.811 | 1.282 | 2588 | 2106 | 3180 |

Table 6.10a. Cod in Division 7e-k. Table of model parameters.

| Parameter name | par | sd(par) | $\exp (\mathrm{par})$ | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: |
| logFpar_0 | -6.920 | 0.050 | 0.001 | 0.001 | 0.001 |
| logFpar_1 | -3.149 | 0.159 | 0.043 | 0.031 | 0.059 |
| logFpar_2 | -2.190 | 0.156 | 0.112 | 0.082 | 0.153 |
| logFpar_3 | -1.929 | 0.156 | 0.145 | 0.106 | 0.198 |
| logSdLogFsta_0 | $-2.480$ | 0.465 | 0.084 | 0.033 | 0.212 |
| $\operatorname{logSdLogN}$ | -0.079 | 0.132 | 0.924 | 0.710 | 1.202 |
| logSdLogN_1 | -2.012 | 0.396 | 0.134 | 0.061 | 0.295 |
| logSdLogObs_0 | -0.707 | 0.214 | 0.493 | 0.322 | 0.756 |
| logSdLogObs_1 | -1.129 | 0.251 | 0.323 | 0.196 | 0.534 |
| logSdLogObs_2 | -1.235 | 0.110 | 0.291 | 0.233 | 0.362 |
| logSdLogObs_3 | -1.850 | 0.218 | 0.157 | 0.102 | 0.243 |
| logSdLogObs_4 | -0.449 | 0.139 | 0.638 | 0.483 | 0.843 |
| transfIRARdist_0 | -0.754 | 0.447 | 0.470 | 0.193 | 1.149 |
| itrans_rho_0 | 1.849 | 0.706 | 6.356 | 1.548 | 26.092 |

Table 6.10b. Cod in Division 7e-k. Model fitting.

| Model | $\log (\mathrm{L})$ | \#par | AIC |
| :--- | :--- | :--- | :--- |
| Current | -221.51 | 14 | 471.01 |
| base | -209.68 | 14 | 447.37 |

Table 6.11. Cod Division 7e-k. Short-term forecast assumption.

| Variable | Value | Notes |
| :--- | :---: | :--- |
| F ages 2-5 (2022) | 1.139 | $\mathrm{~F}_{\text {sq }}=\mathrm{F}_{\text {average }}$ (2019-2021) |
| SSB (2023) | 992 | Fishing at $\mathrm{F}_{\text {sq; }}$ in tonnes. |
| Recruitment age 1 (2022-2023) | 1305,1305 | Median from resampled (2015-2021); in thousands |
| Total catch (2022) | 1220 | Fishing at Fsq, in tonnes. |

Table 6.12. Cod in Division 7e-k. Reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 5800 | $\mathrm{B}_{\mathrm{pa}}$; in tonnes | ICES (2020a) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.29 | Segmented regression with $\mathrm{Blim}_{\text {( }}$ (EqSim) | ICES (2020a) |
| Precautionary approach | $\mathrm{Blim}_{\text {lim }}$ | 4200 | $\mathrm{B}_{\text {loss, }}$ lowest observed SSB from which there has been some recovery (2005) rounded value; in tonnes | ICES (2020a) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 5800 | $\mathrm{B}_{\lim } \times 1.4$; in tonnes | ICES (2020a) |
|  | $\mathrm{F}_{\text {lim }}$ | 1.13 | Segmented regression with $\mathrm{Blim}_{\text {l }}$ (EqSim) | ICES (2020a) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.77 | $\mathrm{F}_{\mathrm{p} 0.5}$; the F that leads to $\mathrm{SSB} \geq \mathrm{Bl}_{\text {lim }}$ with $95 \%$ probability | ICES (2020a) |
| Management plan | MAP <br> MSY $B_{\text {trigger }}$ | 5800 | MSY $\mathrm{B}_{\text {trigger }}$; in tonnes | $\begin{aligned} & \text { EU (2019), ICES } \\ & \text { (2020a) } \end{aligned}$ |
|  | MAP Blim | 4200 | $\mathrm{B}_{\text {lim }}$; in tonnes | $\begin{aligned} & \text { EU (2019), ICES } \\ & \text { (2020a) } \end{aligned}$ |
|  | MAP $\mathrm{F}_{\text {MSY }}$ | 0.29 | $\mathrm{F}_{\text {MSY }}$ | $\begin{aligned} & \text { EU (2019), ICES } \\ & (2020 a) \end{aligned}$ |
|  | MAP range $\mathrm{F}_{\text {lower }}$ | 0.17 | Consistent with ranges resulting in no more than 5\% reduction in long-term yield compared with MSY | $\begin{aligned} & \text { EU (2019), ICES } \\ & \text { (2020a) } \end{aligned}$ |
|  | MAP range $\mathrm{F}_{\text {upper }}$ | 0.41 | Consistent with ranges resulting in no more than 5\% reduction in long-term yield compared with MSY | $\begin{aligned} & \text { EU (2019), ICES } \\ & \text { (2020a) } \end{aligned}$ |

Table 6.13. Stochastic Short-term forecast.
$F_{M S Y}$

| Year | Fbarme- <br> dian | $F_{\text {barlow }}$ | Fbar <br> high | Rec <br> me- <br> dian | Rec <br> low | Rec <br> high | SSB me- <br> dian | SSB <br> low | SSB <br> high | Catch me- <br> dian | Catch <br> low | Catch <br> high |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 1.047 | 0.816 | 1.344 | 925 | 471 | 1812 | 1346 | 1052 | 1724 | 1333 | 1065 | 1677 |
| 2022 | 1.139 | 0.847 | 1.535 | 1305 | 666 | 3502 | 1196 | 732 | 1879 | 1220 | 833 | 1805 |
| 2023 | 0.290 | 0.207 | 0.406 | 1305 | 666 | 3502 | 992 | 490 | 1909 | 406 | 247 | 759 |
| 2024 | 0.290 | 0.199 | 0.420 | 1305 | 666 | 3502 | 2042 | 1040 | 4137 | 740 | 429 | 1424 |

Basis for the advice $F=0$

| Year | $F_{\text {barme- }}$ <br> dian | $F_{\text {barlow }}$ | Fbar <br> $\mathbf{h}_{\text {igh }}$ | Rec <br> me- <br> dian | Rec <br> low | Rec <br> high | SSB me- <br> dian | SSB <br> low | SSB <br> high | Catch me- <br> dian | Catch <br> low | Catch <br> high |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 1.047 | 0.816 | 1.344 | 925 | 471 | 1812 | 1346 | 1052 | 1724 | 1333 | 1065 | 1677 |
| 2022 | 1.139 | 0.847 | 1.535 | 1305 | 666 | 3502 | 1196 | 732 | 1879 | 1220 | 833 | 1805 |
| 2023 | 0.000 | 0.000 | 0.000 | 1305 | 666 | 3502 | 992 | 490 | 1909 | 0 | 0 | 0 |
| 2024 | 0.000 | 0.000 | 0.000 | 1305 | 666 | 3502 | 2649 | 1419 | 5309 | 0 | 0 | 0 |

Other scenarios

F status quo then $\mathrm{F}_{\text {msy }} \mathrm{HCR}$

| Year | $F_{\text {barmedian }}$ | $F_{\text {barlow }}$ | $F_{\text {bar high }}$ | Rec median | Rec low | Rec high | SSB median | SSB low | SSB high | Catch median | Catch low | Catch high |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 1.047 | 0.816 | 1.344 | 925 | 471 | 1812 | 1346 | 1052 | 1724 | 1333 | 1065 | 1677 |
| 2022 | 1.139 | 0.847 | 1.535 | 1305 | 666 | 3502 | 1196 | 732 | 1879 | 1220 | 833 |  |
| 2023 | 0.050 | 0.035 | 0.069 | 1305 | 666 | 3502 | 992 | 490 | 1909 | 77 | 47 |  |
| 2024 | 0.290 | 0.199 | 0.420 | 1305 | 666 | 3502 | 2531 | 1345 | 5088 | 890 | 146 |  |

F status quo then FMSY lower HCR

| Year | $F_{\text {barmedian }}$ | $F_{\text {barlow }}$ | $F_{\text {bar high }}$ | Rec median | Rec low | Rec high | SSB median | SSB low | SSB high | Catch median | Catch low | Catch high |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 1.047 | 0.816 | 1.344 | 925 | 471 | 1812 | 1346 | 1052 | 1724 | 1333 | 1065 | 1677 |
| 2022 | 1.139 | 0.847 | 1.535 | 1305 | 666 | 3502 | 1196 | 732 | 1879 | 1220 | 833 | 1805 |
| 2023 | 0.029 | 0.021 | 0.041 | 1305 | 666 | 3502 | 992 | 490 | 1909 | 46 | 86 |  |
| 2024 | 0.290 | 0.199 | 0.420 | 1305 | 666 | 3502 | 2579 | 1375 | 5183 | 904 | 521 | 1766 |

F status quo then FmsY upper HCR

| Year | $F_{\text {barmedian }}$ | $F_{\text {barlow }}$ | $F_{\text {bar high }}$ | Rec median | Rec low | Rec high | SSB median | SSB low | SSB high | Catch median | Catch low | Catch high |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 1.047 | 0.816 | 1.344 | 925 | 471 | 1812 | 1346 | 1052 | 1724 | 1333 | 1065 | 1677 |
| 2022 | 1.139 | 0.847 | 1.535 | 1305 | 666 | 3502 | 1196 | 732 | 1879 | 1220 | 833 | 1805 |
| 2023 | 0.070 | 0.050 | 0.098 | 1305 | 666 | 3502 | 992 | 490 | 1909 | 108 | 66 |  |
| 2024 | 0.290 | 0.199 | 0.420 | 1305 | 666 | 3502 | 2485 | 1316 | 4998 | 876 | 505 |  |

Stable SSB

| Year | $F_{\text {barmedian }}$ | $F_{\text {barlow }}$ | $F_{\text {bar high }}$ | Rec median | Rec low | Rec high | SSB median | SSB low | SSB high | Catch median | Catch low | Catch high |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 1.047 | 0.816 | 1.344 | 925 | 471 | 1812 | 1346 | 1052 | 1724 | 1333 | 1065 | 1677 |
| 2022 | 1.139 | 0.847 | 1.535 | 1305 | 666 | 3502 | 1196 | 732 | 1879 | 1220 | 833 | 1805 |
| 2023 | 1.116 | 0.796 | 1.563 | 1305 | 666 | 3502 | 992 | 490 | 1909 | 1136 | 681 |  |
| 2024 | 1.145 | 0.787 | 1.660 | 1305 | 666 | 3502 | 992 | 426 | 2173 | 1195 | 2078 |  |

$F_{2022}$

| Year | $F_{\text {barmedian }}$ | $F_{\text {barlow }}$ | $F_{\text {bar high }}$ | Rec median | Rec low | Rec high | SSB median | SSB low | SSB high | Catch median | Catch low | Catch high |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 1.047 | 0.816 | 1.344 | 925 | 471 | 1812 | 1346 | 1052 | 1724 | 1333 | 1065 | 1677 |
| 2022 | 1.139 | 0.847 | 1.535 | 1305 | 666 | 3502 | 1196 | 732 | 1879 | 1220 | 833 | 1805 |
| 2023 | 1.139 | 0.813 | 1.596 | 1305 | 666 | 3502 | 992 | 490 | 1909 | 1151 | 689 | 2104 |
| 2024 | 1.139 | 0.783 | 1.652 | 1305 | 666 | 3502 | 972 | 416 | 2135 | 1175 | 654 | 2192 |

$\mathrm{F}_{\mathrm{pa}}$

| Year | $F_{\text {barmedian }}$ | $F_{\text {barlow }}$ | $F_{\text {bar high }}$ | Rec median | Rec low | Rec high | SSB median | SSB low | SSB high | Catch median | Catch low | Catch high |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2021 | 1.047 | 0.816 | 1.344 | 925 | 471 | 1812 | 1346 | 1052 | 1724 | 1333 | 1065 |  |
| 2022 | 1.139 | 0.847 | 1.535 | 1305 | 666 | 3502 | 1196 | 732 | 1879 | 1220 | 833 |  |
| 2023 | 0.770 | 0.549 | 1.078 | 1305 | 666 | 3502 | 992 | 490 | 1909 | 890 | 538 |  |
| 2024 | 0.770 | 0.530 | 1.116 | 1305 | 666 | 3502 | 1336 | 618 | 2813 | 1148 | 1637 |  |



Blim - Not archivable.
$B_{p a}$, MSYBtrigger, not archivable.
Table 6.14. Catch option table.

| Basis | Total catch (2023) | $F_{\text {total }}(2023)$ | SSB (2024) | \% SSB change * |
| :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |
| MSY and precautionary considerations: $F=0$ | 0 | 0 | 2649 | 167 |
| Other scenarios |  |  |  |  |
| $\mathrm{F}_{\text {MSY }} \times \mathrm{SSB}_{2023} / \mathrm{MSY}^{\text {B }}$ trigger | 77 | 0.050 | 2531 | 155 |
| $\mathrm{F}_{\text {MSY lower }} \times \mathrm{SSB}_{2023} / \mathrm{MSY}^{\text {B }}$ trigger | 46 | 0.029 | 2579 | 160 |
| $\mathrm{F}_{\text {MSY upper }} \times \mathrm{SSB}_{2023} / \mathrm{MSY} \mathrm{B}_{\text {trigger }}$ | 108 | 0.070 | 2485 | 150 |
| $\mathrm{F}=\mathrm{F}_{\text {MSY }}$ | 406 | 0.290 | 2042 | 106 |
| $\mathrm{F}=0$ | 0 | 0 | 2649 | 167 |
| $\mathrm{F}=\mathrm{F}_{\text {lim }}$ | 1145 | 1.130 | 980 | -1.2 |
| $\mathrm{F}=\mathrm{F}_{\mathrm{pa}}$ | 890 | 0.770 | 1336 | 35 |
| $\mathrm{SSB}_{2024}=\mathrm{Bl}_{\text {lim }}$ |  |  |  |  |
| $\mathrm{SSB}_{2024}=\mathrm{Bpa}=\mathrm{MSY} \mathrm{B}_{\text {trigger }}$ |  |  |  |  |
| $F=F_{2022}$ | 1151 | 1.139 | 972 | -2 |
| $\mathrm{SSB}_{2024}=$ SSB $_{2023}$ | 1136 | 1.116 | 992 | 0 |

Table 6.15.Cod in Division 7e-k. Forecast (a) yield in 2023 and (b) SSB in 2024.

| recruitment | val | type | Prop | Age |
| :---: | :---: | :---: | :---: | :---: |
| 2023 | 0,00024668 | 2023 Catch | 17,194287 | 1 |
| 2022 | 0,0006068 | 2023 Catch | 42,2951259 | 2 |
| 2021 | 0,00024928 | 2023 Catch | 17,3752029 | 3 |
| 2020 | 0,00014912 | 2023 Catch | 10,3938527 | 4 |
| 2019 | 0,00015426 | 2023 Catch | 10,7518776 | 5 |
| 2018 | 2,24E-05 | 2023 Catch | 1,55794101 | 6 |
| 2017 | 6,19E-06 | 2023 Catch | 0,43171295 | 7 |
| 2024 | 0 | 2024 SSB | 0 | 1 |
| 2023 | 651,38811 | 2024 SSB | 25,0285744 | 2 |
| 2022 | 1188,47341 | 2024 SSB | 45,6652411 | 3 |
| 2021 | 375,213077 | 2024 SSB | 14,4169785 | 4 |
| 2020 | 186,541001 | 2024 SSB | 7,16754763 | 5 |
| 2019 | 176,350869 | 2024 SSB | 6,77600769 | 6 |
| 2018 | 24,6112966 | 2024 SSB | 0,94565077 | 7 |



Figure 6.1. Cod in Division 7e-k. Historical landings (in tonnes) by country. Revised at WKCELTIC 2020.


Figure 6.2. Cod in Division 7e-k. Catches volume in Tonnes (i.e.landings and discards) by area and country.


Figure 6.3. Cod in Division 7e-k. Proportion of the catches per métier (Level 6).


Figure 6.4. Cod in Division 7e-k. Discard proportion per métier and season.


Figure 6.5. Cod in Division 7e-k. Allocation procedure.


Figure 6.6a. Cod in Division 7e-k. Time-series of (a) LPUE and (b) fishing effort for the French fleets. Units: LPUE in kg/day and fishing effort in days fished.


Figure 6.6b. Cod in Division 7e-k. Time-series of (a) LPUE and (b) fishing effort for the Irish fleets. Units: LPUE in kg/day fished and Effort in 000s hours fished.


Figure 6.6c. Cod in Division 7e-k. Time-series of LPUE and fishing effort for the UK fleets. Units: LPUE in kg/day and fishing effort in days fished.



Figure 6.7. Cod in Division 7e-k. Fits of the tuning indices used in the assessment. Commercial tuning fleet corresponds to French OTDEF Q2+3+4 as biomass index. The survey index is a combined index based on both French IR-GFS and FR Evhoe Q4 data where mean number-at-age are modelled using VAST.


Figure 6.8. Cod in Division 7e-k. Final assessment. Residuals.


Figure 6.9. Cod in Division 7e-k. Fishing mortality.


Figure 6.10. Cod in Division 7e-k. Final SAM outputs. Catch proportion-at-age. Age $\mathbf{0}$ are not included in the assessment.




Figure 6.11. Cod in Division 7e-k. Final SAM outputs. SSB, F, R and catches estimates.




Figure 6.12a. Cod in Division 7e-k. Final SAM. Retrospective plots.




Figure 6.12b. Cod in Division 7e-k. Final SAM. Comparison between runs (runs with the two tuning indices, with only the survey index and with only the commercial tuning index).


Figure 6.13. Cod in Division 7e-k. Stock-recruitment plots and yield per recruit information.


Figure 6.14. Cod in Division 7e-k. Forecast (a) yield in 2023 and (b) SSB in 2024.

## 7 Haddock in Division 6.b (Rockall)

### 7.1 Introduction

In previous years an age-structured assessment model has been used to provide advice as a category 1 stock. This year, methods to update the previously utilised survey index were unavailable and the agreed assessment (and forecast) could not be carried out. Consequently, the assessment this year is based on a new biomass index derived from the Rock-WIBTS-Q3 as an indicator of stock size and a mean catch length indicator as a proxy for fishing pressure (under a category 3 approach). The WG considers this to be a temporary measure until such a time as a benchmark for this stock can be scheduled.

The derivation of the biomass index and the estimation of data/parameters required for the length indicators are described in further detail below.

### 7.2 General

## Advice

ICES advice has been provided on the basis of the MSY approach since 2014. Last year's advice was for catches of no more than 5825 tonnes (for 2022).

## Stock description and management units

The haddock stock at Rockall is considered to be an entirely separate stock from that inhabiting the continental shelf of the British Isles. Since 2004, the EU TAC for haddock in $6 . b$ has been included with Divisions 12 and 14. For details of the earlier management units see the Stock Annex.

## Management applicable to 2021 and 2022

The TAC is set for the UK, EU and international waters of 6b, and international waters of 12 and 14. For 2021 and 2022, the breakdown by country is given below:

TAC 2021

| Species: | Haddock <br> Melanogrammus acglefinus | Zone: $\quad$United Kingdom, Union and international waters <br> of 6 b ; international waters 12 and 14 <br> $(\mathrm{HAD} / 6 \mathrm{~B} 1214)$ |
| :---: | :---: | :---: |
| Belgium | 16 | Analytical TAC |
| Germany | 19 | Article 8(2) of this Regulation applies |
| France | 799 |  |
| Ireland | 570 |  |
| Union | 1404 |  |
| United <br> Kingdom | 6971 |  |
| TAC | 8375 |  |

TAC 2022

| Species: | Haddock <br> Melanogrammus acglefinus |  | Zone: | United Kingdom, Union and international waters of 6 b ; international waters of 12 and 14 (HAD/6B1214) |
| :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 12 | Analy |  |
| Germany |  | 12 | Article | this Regulation applies |
| France |  | 542 |  |  |
| Ireland |  | 385 |  |  |
| Union |  | 951 |  |  |
| United Kingdom |  | 4874 |  |  |
| TAC |  | 5825 |  |  |

The ICES advice, agreed TAC for EU waters, and WG estimates of landings during 2002-2021 are summarised below. All values are in tonnes.

| YEAR | Predicted catch corresp. to advice | Predicted landings corresp. to advice\# | BASIS | AGREED <br> TAC ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: |
| 2002 | <1300 |  | Reduce F below 0.2 |  |
| 2003 | - |  | Lowest possible F |  |
| 2004 | - |  | Lowest possible $\mathrm{F}^{\text {b }}$ | 702 |
| 2005 | - |  | Lowest possible F ${ }^{\text {b }}$ | 702 |
| 2006 | - |  | Lowest possible $\mathrm{F}^{\text {b }}$ | 597 |
| 2007 | < 7100 |  | Reduce F below FPA ${ }^{\text {b }}$ | 4615 |
| 2008 | < 10640 |  | Keep F below FPA ${ }^{\text {b }}$ | 6916 |
| 2009 |  | < 4300 | No long-term gains in increasing $\mathrm{F}^{\mathrm{b}}$ | 5879 |
| 2010 |  | <3300 | Little gain on the long-term yield by increasing F b | 4997 |
| 2011 |  | <2700 | Reduction in F is needed to keep SSB to above BPA in 2012 | 3748 |
| 2012 |  | <3300 | MSY approach | 3300 |
| 2013 | 0 | 0 | No directed fisheries, minimize bycatch and discards | 990 |
| 2014 | $<1620$ | <0980 | MSY approch | 1210 |
| 2015 | $<4310$ | <2930 | MSY approch | 2580 |
| 2016 | < 3932 | < 3225 | MSY approach | 3225 |
| 2017 | $\leq 4690$ | $\leq 4130$ | MSY approach | 4690 |
| 2018 | $\leq 5163$ |  | MSY approach | 5163 |
| 2019 | $\leq 10469$ |  | MSY approach | 10469 |
| 2020 | $\leq 10472$ |  | MSY approach | 10472 |
| 2021 | $\leq 6239$ |  | MSY approach | 8375 |
| 2022 | $\leq 5825$ |  | MSY approach | 5825 |

a Prior to 2014, the TAC was set for Divisions 6.a and 6.b (plus 5.b, 12 and 14) combined with restrictions on quantity that can be taken in $5 . b$ and 6.a. The quantity shown here is the total area TAC minus the maximum amount which is

```
allowed to be taken from 5.b and 6.a. In 2004, the EU TAC for Division }6\mathrm{ was split and the 6.b TAC for haddock was included with 12 and 14. This value is the TAC for 6.b, 12 and 14.
\({ }^{\text {b }}\) Single-stock boundary and the exploitation of this stock should be conducted in the context of mixed fisheries, protecting stocks outside safe biological limits.
```

The minimum conservation reference size of haddock taken by EU and UK vessels at Rockall is 30 cm . There is no minimum landing size for haddock taken by non-EU/UK vessels within international waters.

In order to protect the pre-recruit stock, the International Waters component of the statistical rectangle 42D5 has been closed for fishing since 2001 and its EU component since 2002 (see the Stock Annex, Section A.3). The protected area (the whole rectangle) is referred to as the Rockall Haddock Box. In 2022, in response to a request for advice from NEAFC on the efficiacy of the closure to protect juvenile haddock, ICES concluded the following:

The Rockall Haddock Box does coincide with areas of high juvenile and adult haddock densities, with high densities also observed outside the box to the northeast. For most years since the closure, haddock densities of age classes 1+ have been higher inside than outside the box. The overall impact of the current closure area on the Rockall haddock stock continues to be difficult to assess

In order to protect cold-water corals, three further areas (North West Rockall, Logachev Mounds and West Rockall Mounds) were closed since January 2007 (see the Stock Annex, Section A.3). A new area to protect cold-water corals (Empress of Britain Banks) was established by the NEAFC in 2007 and 2012.

Since 2009 in NEAFC regulatory area, including international waters of Rockall, a discard ban was established. The fishery for haddock within EU/UK waters is under the landing obligation.

There is no agreed management plan for haddock in this area. Two management strategies (NEAFC and EU MAP) have been assessed to be precautionary. NEAFC requested ICES to evaluate the harvest control rules (HCRs) that use FmsY as a target. ICES concluded that the NEAFC HCRs in the long-term management strategy for Rockall haddock were consistent with the precautionary approach (ICES, 2019a).

The multiannual management plan (MAP) which has been adopted by the EU for this stock (EU, 2019) has not been agreed with UK.

## Fishery in 2021

## Russian fishery in 2021

No information was provided on the Russian fishery at Rockall in 2021. Total official landings of haddock at Rockall by the Russian Federation amounted to 20 t (Table 7.3.1).

## UK fishery in 2021

A small number of larger Scottish demersal trawlers continue to target haddock at Rockall with the fishery largely conducted during the summer months and periods of good weather. Other important target species included anglerfish (Lophius spp.), ling, saithe and megrim. Total Scottish haddock landings in 6 b have declined by over $40 \%$ since 2019. Quota uptake was also relatively low ( $<50 \%$ ) in 2021 which may be related to availability of haddock (and quota) on less distant fishing grounds such as in Divisions 6a and 4a. In 2021 there was a significant increase in the proportion of UK landings taken in international waters (almost $30 \%$ in 2021 compared to an average of $15 \%$ over 2018-2020). (Table 7.3.2).

## Irish fishery in 2021

Reported landings by Irish vessels decreased in 2021 compared to 2020 (Table 7.3.1), although as for UK vessels there was an increase in the proportion of landings being taken in the NEAFC area. Irish vessels used single otter trawls with a mesh size ranging from 100 to 120 mm together with a square mesh panel.

### 7.3 Data

## Landings

Nominal landings as reported to ICES are given in Table 7.3.1 and shown in Figure 7.3.1, along with Working Group estimates of total estimated landings. Revisions to official catch statistics for previous years are also shown in Table 7.3.1. Some data previously submitted are now no longer or only partially available due to national data confidentiality clauses (Ireland in 2018 and 2019). As has been the case for over 10 years, the majority of the reported landings come from the UK (almost $90 \%$ in 2021) with smaller amounts reported by Ireland and the Russian Federation.

Data for the NEAFC area only (Sub-division 6b.1) are also shown in Table 7.3.1 and in Table 7.3.2 by nation. Up to 2019, these area taken from the official landings statistics. For 2020 and 2021, the estimates are a mixture of official landings and landings from Intercatch as landings for the UK are not available split by sub-division in the preliminary official landings data. In some years, Russian landings are reported as being from 27.6b_NK (i.e. unknown sub-division), however it is assumed that all these landings have been taken in the NEAFC area (i.e. sub-division 27.6.b.1). The proportion of the total landings coming from international waters is almost $30 \%$ in 2021, a substantial increase compared to 2019 and 2020.

Anecdotal evidence suggests that misreporting of haddock from Rockall has occurred historically (particularly on fishing trips where vessels fish in both Division 6a and 6b), but a quantitative estimation of the degree of misreporting is not possible.

Landings data submitted to Intercatch are shown in Figure 7.3.2. Russian Federation landings were not submitted by the national data submitter for 2021, however a comparison of historical official landings and ICES estimates suggested the two to be consistent and therefore ICES estimated landings for the Russian Federation were assumed equal to the official reported landings for 2021 (and subsequently imported to Intercatch).

International age composition and mean weight-at-age in the landings were compiled according to the methods described in the Stock Annex. Landings age compositions were allocated to unsampled fleets using a weighted average of all sampled fleets (excluding the Russian fleet which retain all landings on board). The weighting algorithm used is 'Mean weight weighted by num-bers-at-age or length'.

The need to use a data limited approach for the assessment and advice this year meant that length frequency data also had to be processed. Landings length compositions were allocated to unsampled fleets in the same manner as age compositions. Data were processed in InterCatch for 2012 onwards.

## BMS landings

In 2016, Below Minimum Size (BMS) landings (subject EU landings obligation) were negligable at 0.4 t . In 2017 and 2018 BMS landings were not reflected in the catch statistic. 4 t of BMS landings were reported in 2019, 2 t in 2020 and $<1 \mathrm{t}$ in 2021. The assessment includes BMS landings within
total landings (although these unsampled data are allocated age/length compositions from sampled discards).

## Discards

Haddock at Rockall have lower size-at-age than haddock from other areas (Blacker, 1971; Khlivnoy, 2006; Filina, Khlivnoy and Vinnichenko, 2009). Historically, the discard rate was as high between 12 and $75 \%$ by weight according to the results of discards trips (see the Stock Annex). The methods used to reconstruct the historical time-series of discards when sampling data were insufficient or unavailable is described in the Stock Annex.

At the 2019 benchmark, the catch-at-age data from 2012 onwards were re-estimated in InterCatch. The two main fleets (UK(Scotland OTB_DEF_>=120 and Irish OTB_DEF_100-119) are $t$ sampled for both landings and discards. Discard rate allocation to other unsampled fleets consisted of:

- Manually matching annual discards to available quarterly landings by country/fleet (where necessary);
- Using a weighted average discard rate for all unsampled fleets (weighted by CATON) with the exception of the Norwegian longline fleet and the Russian fleet for which discards are both assumed to be zero.

Discards age and length compositions are allocated to unsampled fleets in a similar manner to landings age/length compositions. This process has been conducted annually since the 2019 benchmark.

Figure 7.3 .3 shows estimated landings and discards for 2021 after raising. Scottish landings data are submitted to InterCatch by sub-division with sampling available only for sub-division 6.b. 2 and hence there are raised discards associated with this component of the landings. The final mix of numbers-at-length from sampled and un-sampled landings and sampled and raised (unsampled) discards is given in Figure 7.3.4. The unsampled landings and unsampled raised discards are both associated with the Scottish landings from sub-division 6.b.1.
During the 2019 benchmark meeting, the discards data for 2010 were also revised. The discards were calculated on average discards proportion of total catch but prior to the 2019 benchmark, the discards only for 2010 were calculated on that proportion applied to the landings (not total catch). The benchmark concluded that the previous method applied for 2010 was incorrect, as discards proportion is relative the total catch. Since 2019 in correct assessment, the discards for 2010 were calculated based on the discards numbers in 2009 at-age recalculated using the ratio between total landings in 2009 and 2010 (by numbers).

The WG estimates of total landings and discards by weight are given in Table 7.3.3 and shown in Figure 7.3.5. In recent years, the total discard proportion by weight (Figure 7.3.6) has shown substantial variability but is typically below $20 \%$ ( $21 \%$ in 2021). This is substantially lower than the estimated discard proportion at the start of the time series when the Scottish fleet was generally utilising a smaller mesh size.

Due to the distant nature of the fishery and the fact that there are relatively few vessels/trips making landings, sampling levels for both landings and discards for this stock are relatively poor. Sampling levels did not significantly worsen during the COVID-19 pandemic and the WG considers that sampling levels are adequate for the estimation of catch length-based indicators.

## Age- and length-compositions and mean weights-at-age

Raised landings numbers-at-age and discard numbers-at-age are given in Tables 7.3.4 and 7.3.5 respectively and total catch numbers-at-age in Table 7.3.6. Although these are not required for the data limited approach utilised this year, they are included here for completeness.

The catch-at-length compositions which are used to derive length based indicators for the data limited approach utilised this year are shown in Figure 7.3.7.

Annual mean weights-at-age in landings, discards and catch are given in Tables 7.3.7, 7.3.8 and 7.3.9 and shown in Figure 7.3.8. Mean weights-at-age in the landings (and catch) were relatively stable historically, but show a significant increase since the mid-2000s, particularly at ages 4 and above, as well as showing increased variability. While the variability may be due to low sampling levels, the reason for the trend is not known, but could potentially be due to cohort dependent growth rates although this has not been fully explored.

## Biological

There was no change in natural mortality and maturity compared last year's assessment (see the Stock Annex) although neither are used in the data limited approach employed this year.

Historically, stock weights-at-age were assumed to be equal to the raw catch weights. However, the number of sampled trips for both landings and discards has been low for a number of years and this appears to have led to substantial variability in the mean weight-at-age estimates. For this reason, at the 2019 benchmark, it was agreed that five year rolling average catch mean weights-at-age should be used as stock mean weights-at-age. These are given in Table 7.3.10.

As a check on the trends in the stock mean weights-at-age derived from catch mean weights-atage, mean weights-at-age were also derived from survey data and a comparison made (Figure 7.3.9). Survey mean weights-at-age are calculated by i) estimating an annual length-weight relationship (from the samples in the DATRAS 'CA' data), ii) calculating individual weight-atlength using the length-weight parameters, iii) deriving an annual length frequency -at-age by raising the ALK (from the 'CA' data) to the total length distribution (from 'HL' data), iv) calculating mean weight-at-age using the results of ii) and iii). (Pre 2011 an average weight-length relationship over all available years was used as no individual weight data were recorded in this period).

Although there are missing values in the survey data, values appear reasonably consistent with those from the catch and there is some evidence of an increase since the early 2000s (although more recently, survey estimates appear somewhat lower). For the derivation of the total biomass index used as the stock indicator, the stock mean weights-at-age 1-7 are derived from the catch mean weights as per previous assessment WGs ( 5 year rolling mean) while age 0 stock mean weights-at-age are derived from the survey as these are largely not available from the catch (and in the years they are available, estimates are likely to be impacted by gear selectivity).

## Surveys

The Scottish Rock-IBTS-Q3 survey is the only survey available for this stock. The survey is coordinated by IBTS and described further in the IBTS reports, the Stock Annex and the 2019 benchmark report (ICES, 2019).

The survey originally began in the 1990s, but in 2011, a number of changes were made: the survey groundgear was changed from GOV-C to GOV-D and the survey changed from a fixed station design to random stratified covering a greater depth range than previously. Studies
conducted in 2006 and 2009, comparing the catchability of the two groundgears (WD 5 in ICES, 2012a) suggested no significant differences in the catch rate of haddock. The assessment WG in 2012 (ICES, 2012b) proposed an approach which was agreed at the 2019 benchmark, whereby only the subset of survey stations occurring within the depth range of the original (pre 2011) survey were included in the post 2011 index calculation. This allowed the survey to be treated as a continuous time series in the assessment (known as the 'standard index'). This year, the WG was unable to recreate the calculation of the 'standard index' and therefore the 2021 index value could not be calculated according to this approach. Therefore, a new index was calculated according to the survey design.

In the new index calculation (2011 onwards), numbers at length per haul are standardised to numbers per hour towing (LFD). An ALK is calculated for each of the four strata and then this ALK is applied to the LFDs from each of the hauls separately to produce age frequencies for each haul. Finally, for each stratum, the age frequencies are summed and the values divided by the number of valid hauls to provide numbers at age per hour. This procedure can be summarised as

$$
\operatorname{CPUE}_{i, a}=\frac{\sum_{h=1}^{H_{i}} \sum_{l=l_{\min }}^{l=l_{\max }} N_{i, a, l, h}}{H_{i}}
$$

where $N_{i, a, l, h}$ is the number of fish at age $a$ and length $l$ caught during haul $h, H_{i}$ is the number of valid hauls in stratum $i$ and $C P U E_{i, a}$ is the catch per unit effort of fish at age $a$ in stratum $i$.

For each age, the age frequency for each stratum is raised by the stratum area. These raised frequencies are then summed and the result divided by the total area in the assessment region. The final index value for each age is given by

$$
I_{a}=\frac{\sum_{i=1}^{S} C P U E_{i, a} A_{i}}{\sum_{i=1}^{S} A_{i}}
$$

where $A_{i}=$ area $\left(\mathrm{m}^{2}\right)$ of stratum $i$ and $S=$ number of strata. Survey variance estimates are calculated in a similar manner.

The old index (pre-survey change in 2011) is given in Table 7.3.11. The new index (estimate and variance) is given in Table 7.3.12. The previously used continuous 'standard index' is not presented but can be found in last year's WG report (ICES, 2021).

Plots of survey log cpue at age by cohort are shown in Figure 7.3.10 and comparative scatterplots of log index at-age are shown in Figure 7.3.11. The survey shows good internal consistency in tracking of year-class strength through time.

In Figure 7.3.12 compares the new (2011 onwards) index with that used in previous assessment WGs mean standardised over the common time period. The two indices show small differences, mostly confined to age 0 and the older ages, where the new index appears to be higher for the strong year classes.

## Commercial Effort, Lpue and Cpue

Commercial effort data have previously been provided for Scottish, Irish and Russian fleets, along with lpue and cpue data respectively. These data have not been updated in recently years
and have not been used for assessment purposes and are therefore not included here (See 2020 WG report for data).

### 7.4 Derivation of stock indicators

## Survey index

The biomass index for use in the data limited approach to advice was derived as the sum of products of the survey numbers-at-age (Table 7.3.12) and stock mean weights-at-age (Table 7.3.10). The index is shown in Figure 7.3.13 (and Table 7.3.13) in comparison and broken down by age class in Figure 7.3.14. The survey biomass index increases from its lowest value at the start of the time series and shows a general decline since 2018. The SSB from the 2021 assessment also shows an increase although this does not begin until after 2014 when the individuals are mature (knife edge maturity at age 3 assumed in the assessment).

## Growth parameter estimation

Rockall haddock is a poorly studied stock and there is limited published information on individual growth. Fishbase quote a value of $L_{\infty}=43.8 \mathrm{~cm}$ and $\mathrm{k}=0.269$ for von Bertalanffy growth parameters for 'Rockall Island' based on Blacker (1971), although the latter publication does not appear to contain these values or the data source from which they are derived, but indicates that this stock is slower growing than others. Fishbase also contains a range of estimates from other haddock stocks around the UK (Figure 7.3.16). There are multiple L $\infty$ values for North Sea haddock which range from 48 cm to 74 cm and the diversity of values may be associated with cohort dependent growth which is believed to occur in haddock stocks (due to varying cohort size and associated resources).

Given that the Scottish Q3 Rockall haddock survey data contains a substantial amount of length and age sampling, von Bertalanffy growth parameters were estimated using these data. Although it is acknowledged that deriving growth parameters based on length-stratified age samples (the 'CA' data in Datras) results in biased estimates (Perreault et al., 2020) there appears to be no agreed approach as to how best to account for the overall length frequency data in the calculation. The approach taken here was to calculate the length given age (using ' $\mathrm{CA}^{\prime}$ ' and ' HL ' data from Datras) and use a weighted nls regression with each length weighted according to the proportion at length given age. (such that proportions sum to 1 for each age). All survey data from 1999 onwards were included and vB parameters were estimated for each cohort separately for cohorts with more than 6 age classes. Figure 7.3 .15 shows the fitted vB curves and length-age data points (although the weighting of each point is not shown). It is clear from this figure that when there are no age zero data points, the k value cannot be estimated correctly (much too small) resulting in an almost linear vB curve with very high $\mathrm{L}_{\infty}$. However, it is also clear that there is significant variability in estimates between cohorts. The WG therefore agreed to take the median $\mathrm{L}_{\infty}(55 \mathrm{~cm})$ and $\mathrm{k}(0.24)$ values forward for use in the data limited approach. These values are compared to those in Fishbase in Figure 7.3.16. The value is around the middle of the other estimated values while $L_{\infty}$ appears to be towards the lower end of the range of values (but higher than the previously quoted value for Rockall haddock).

## Length-based indicator

Initially annual estimates were made for $L c$ and $L f=M$ (assuming $L i n f=55$ ) and are shown in Table 7.3.14. The annual estimate of Lc is calculated as the lower boundary of the length class which has length frequency at half the maximum (i.e the half the frequency at the modal length). The last 5 years (2017-2021) was taken as the period over which to calculate Lc for use in the lengthbased indicator (and associated reference point) calculation. During this period, catch length frequency data do not appear to be too noisy (Figure 7.3.7) and the stock consists of the full range of lengths (i.e. recruitment has been reasonable over a number of years). Averaged over 20172021, $\mathrm{Lc}=31.6$ and the resulting length-based indicator reference point, $\mathrm{LF}=\mathrm{M}=37.4 \mathrm{~cm}$. The mean length of individuals $>\operatorname{Lc}(31.6 \mathrm{~cm})$ in the catch was then calculated on an annual basis. The resulting mean length estimates (and LBI i.e. values relative to the $L_{\mathrm{F}=\mathrm{m}}$ reference point) are given in Table 7.3.15. LBI values are greater than one for the whole time series for which they are available which suggests that the stock is being fished below FMSY.

The development of the indicators in relation to their reference points is shown in Figure 7.3.17.

### 7.5 Derivation of data limited reference points

The only reference points used in the assessment this year are those associated with the data limited approach. MSY Btrigger is defined from the lowest observed survey index ( $66.764 \mathrm{~kg} \mathrm{hr}^{-}$ ${ }^{1}$ in 2011). The fishing pressure proxy indicator (length-based indicator) is always given in relative terms and therefore the indicator reference point value is 1 . The MSY proxy reference length

| Framework | Reference point | Value | Technical basis |
| :---: | :---: | :---: | :---: |
| $\begin{array}{ll} \text { MSY } & \text { ap- } \\ \text { proach } \end{array}$ | MSY Btrigger | 93.5 | Biomass index trigger value (Itrigger), defined as Itrigger $=I_{\text {loss }} \times 1.4$, where $I_{\text {loss }}$ is the lowest observed historical biomass index value from 2011. In kg per hour. |
|  | FMSY proxy | 1 | Lmean/LF=m; Mean catch length divided by MSY proxy reference length ( $\mathrm{L}=\mathrm{m}=37.4 \mathrm{~cm}$ ). |

### 7.6 Application of the advice rule

The ICES data limited approach for category 3 stocks was applied. According to the new WKLIFE approach, given that this stock has a von Bertalanffy k value between 0.2 and 0.32 , the ' rfbm ' rule with $\mathrm{m}=0.9$ is applied. The WG considered that recent catches have not been too different compared to the advice and therefore the starting point for the application of the rule was last year's advice. This rule uses the application of a multiplier based on the recent trend in a stock biomass index (r), the fishing pressure proxy (f), a biomass safeguard (b) and a precautionary multiplier (m). The stock has declined by $38 \%$ in recent years, however, it is still well above the Itrigger value and therefore the biomass safeguard is 1 . In this case the stability clause is applied to limit the decline in advice to $30 \%$.

| Previous catch advice Ay | 5825 tonnes |
| :---: | :---: |
| Stock biomass trend |  |
| Index A (2020, 2021) | $401 \mathrm{~kg} \mathrm{hr}^{-1}$ |
| Index B (2017, 2018, 2019) | $646 \mathrm{~kg} \mathrm{hr}^{-1}$ |
| r: Index ratio (A/B) | 0.62 |
| Fishing pressure proxy |  |
| Mean catch length (Lmean $=$ L2021) | 39.0 cm |
| MSY proxy length ( $\mathrm{LF}=\mathrm{M}$ ) | 37.4 cm |
| f: multiplier for relative mean length in catches ( $L_{\text {mean }} / L_{\mathrm{F}}=\mathrm{M} 2020$ ) | 1.04 |
| Biomass safeguard |  |
| Last index value (I2021) | $446 \mathrm{~kg} \mathrm{hr}^{-1}$ |
| Index trigger value ( $\mathrm{Itrigger}=\mathrm{I}$ Ioss $\times 1.4$ ) | $93.5 \mathrm{~kg} \mathrm{hr}^{-1}$ |
| b: multiplier for index relative to trigger min\{I202//Itrigger, 1\} | 1 |
| Precautionary multiplier to maintain biomass above Blim with 95\% probability |  |
| m : multiplier (generic multiplier based on life history) | 0.90 |
| RFB calculation** | 3378 tonnes |
| Stability clause ( $+20 \% /-30 \%$ compared to Ay, only applied if Applied b $\geq 1$ ) | 0.70 |
| Discard rate | 9\% |
| Catch advice for 2023 and 2024 (Ay $\times$ stability clause) | 4078 tonnes |
| Projected landings corresponding to advice*** | 3704 tonnes |
| \% advice change^ | -30\% |

### 7.7 Management plans

In 2011 and 2012 in accordance with the conclusions of the 2010-2011, Annual Meeting of the NEAFC, a delegation from the RF and EU considered a management plan. In light of ICES suggestions, the necessary adjustments required to draft a plan were considered. The revised proposal for a harvest control component of a long-term management plan for haddock at Rockall was forwarded to NEAFC for approval at the 2012 Annual Meeting. ICES was requested to evaluate the EU-Russia proposal for the harvest control component of the management plan for Rockall haddock and to evaluate the proposal of protection of juvenile Rockall haddock. The management plan states total catch should not exceed the established TAC and includes measures to record and minimise discards.

ICES evaluated a new HCR proposal for Rockall haddock between RF and EU nations in August 2013 (ICES, 2013) and found that a maximum F of 0.2 was required in the HCR to ensure consistency with the precautionary approach, under the low recruitment conditions observed since 2004.

The NEAFC regulatory area (RA) established a ban on discards. Measures to reduce discards for the stock distribution area were required. The remainder of the management plan for this species is considered to be suitable and has been agreed by the Contracting Parties (NEAFC, 2015).

In 2017, NEAFC requested ICES to evaluate the harvest control component and to consider whether the plan is consistent with the precautionary approach required to provide sustainable harvesting of the stock.

In 2019, ICES evaluated the harvest control rules (HCRs) proposed for Rockall haddock and advised that they are considered precautionary in the short, medium, and long term under the assumption of intermediate levels of productivity.

The HCRs with TAC constraint rule (a) in the request are precautionary in the long term under all scenarios, except those with very low recruitment. If recruitment is low (as observed between 2007 and 2012) over a long time frame, without sporadic recruitment peaks, none of the HCRs
are precautionary in the long term. TAC constraint rule (a) generally leads to lower probability of $\mathrm{SSB}<\mathrm{B}_{\text {lim }}$ than the constraint rule (b), both in the short and long term.

### 7.8 Recommendation for next benchmark

The WG was unable to recreate the survey index used in previous assessments due to the lack of availability of code/spreadsheet. Therefore the agreed (at 2019 benchmark) category 1 assessment could not be conducted and advice had to be provided on the basis of a category 3 stock. Given that the previous category 1 assessment was not rejected and that survey and catch data are considered adequate for an analytical assessment, the WG agreed that this stock should be benchmarked as soon as possible with the following issues.

Type Problem/Aim Work Required | Data Re- |
| :---: |
| quired |

Tuning Calculation of previously used index (de-Develop a modelled index in-Scottish Q3 series rived from Scottish Q3 survey) cannot becluding relevant explanatorysurvey data reproduced due to non availability ofvariables. Explore internal con-(available in code/spreadsheet. Need to agree a newsistency. Comparison with pre-Datras) index for use in the assessment. Due toviously used index \& stratified survey design change in 2011 may re-mean estimates (from MSS quire a break in the index or potentiallywork up). Consider sensitivity modelled index to account for this. of assessment results/quality of
assessment for alternative indi-
ces.
Biologi- Stock mean weights are currently as-Explore catch sampling data forCatch mean cal pa-sumed rolling average of catch meanmean sizes (consistency acrossweights at rameters weights which have shown significantsamples/fleets etc). Use surveyage data. increases (particularly older ages) anddata to derive mean weights atSurvey data variability in recent years. which contrib-age for comparison. Other op-(in Datras). utes to the increase in stock size. Nottions?
clear if this is a low sampling issue or a
real change (cohort effect?).
Biologi- Maturity - currently assumption of knifeReview available data/litera-Maturity cal pa-edge at age 3. Limited data suggeststure. Comparison to otherdata - very rameters some maturity at age 2 . Considerstocks. limited due whether current assumption remains ap- to timing of propriate. current survey (no Q1 survey).

Biologi- Natural mortality. Currently fixed overExplore alternative approachesLife history cal pa-all ages. Last benchmark suggested ex-for deriving natural mortalityparameters. rameters ploring potential for age-dependent Mvalues. based on life-history/mean weights.

[^2]| Type | Problem/Aim | Data Re- <br> quired |
| :---: | :---: | :---: |

assessment method which can account for this uncertainty is required (e.g. SAM).

Biologi- If a new analytical assessment is agreed,Re-estimation of biological refAssessment cal refer-biological ref pts will need to be re-esti-pts. model outence mated
puts. points

Discards Very poor sampling of the fishery. SomeCheck WKROCK data call historical discard data (UK \& EU fleets)(2019), contents of Intercatch \& have been estimated using a combinationwith national data submitters. If of survey data \& selectivity/discardadditional data likely to be ogives derived from years with samples.available, then data call, otherUnlikely to be able to revisit or make re-wise, no data call. visions due to unavailability of raw data. Not clear if all recent data have been uploaded to Intercatch. (Currently 2012 onwards processed in Intercatch).
1.

### 7.9 Management considerations

The fishery for haddock at Rockall is partly in international waters and can therefore be exploited by non UK and EU vessels. The agreed TAC applies only to UK and EU vessels.

Previously, advice for this stock was given following ICES MSY approach based on a Category 1 stock assessment; the applied method gives advice following ICES MSY approach for data limited stocks using the empirical rfb rule (Fischer et al., 2021).

The basis for the advice (rfb rule) has a biennial advice interval and so advice is provided for 2 years (ICES 2022).

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Table 7.3.1. Nominal landings (tonnes) of haddock in Division 6.b, as officially reported to ICES.


* Preliminary official landings.
**Official landings except 2020 \& 2021 which include ICES estimates.
${ }^{\wedge}$ Includes the total Russian catch.
$\wedge \wedge$ Including below minimum size (BMS) landings.
$+<0.5$ tonnes.
\# Updated in 2022.
\#\#Incomplete: part of the data being unavailable under data confidentiality clauses.
n/a = Not available.

Table 7.3.2. Haddock 27.6b. Landings from the NEAFC area (Subdivision 27.6b.1). (Mixture of official landings and ICES estimates - see Section 7.3 for explanation).

| $\begin{aligned} & \frac{1}{\pi} \\ & \underset{\sim}{\sim} \end{aligned}$ |  |  |  | $\begin{aligned} & \text { 믗 } \\ & \underline{\pi} \\ & \underline{\underline{N}} \end{aligned}$ |  | $\begin{aligned} & \text { त } \\ & \frac{3}{0} \\ & 2 \end{aligned}$ |  |  |  | $\begin{aligned} & \cdot \frac{\subseteq}{0} \\ & \text { in } \end{aligned}$ |  | $\begin{aligned} & \overline{\mathrm{T}} \\ & \stackrel{y}{0} \\ & \stackrel{y}{\mathrm{y}} \end{aligned}$ |  | $\begin{aligned} & \bar{\sim} \\ & \stackrel{0}{0} \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 |  |  | 0 |  | 2.2 |  | 4.96 |  | 1 |  | 0 |  | 24.6 |  | 32.8 |
| 2013 |  |  | 0 |  | 4.5 |  | 31.4 |  | 4 |  | 0 |  | 107.2 |  | 147.1 |
| 2014 |  |  | 0 |  | 5.85 |  | 28.9 |  | 388 |  | 0 |  | 0 |  | 422.7 |
| 2015 |  |  | 0 |  | 6.4 |  | 38.6 |  | 136 |  | 0 |  | 59.9 |  | 240.9 |
| 2016 |  |  | 0 |  | 5.2 |  | 47.9 |  | - |  | 0 |  | 511.8 |  | 564.9 |
| 2017 |  |  | 0 |  | 19.9 |  | 7.3 |  | 153 |  | 0 |  | 535.1 |  | 715.3 |
| 2018 |  |  | 0 |  | ** |  | 9.9 |  | - |  | 0 |  | 772.6 |  | 782.5** |
| 2019 |  |  | 1.4 |  | 3.8 |  | 7.3 |  | 245 |  | 0.51 |  | 550.4 |  | 808.5 |
| 2020* |  |  |  |  | 15.3 |  | 11.6 |  | 133 |  |  |  | 584.6 |  | 744.6 |
| 2021* |  | 0.2 | 0.012 |  | 140.9 |  |  |  | 20 |  | 0 |  | 1021.6 |  | 1182.7 |

*Preliminary
**Incomplete: part of the data being unavailable under data confidentiality clauses.

Table 7.3.3. Haddock in 6.b. ICES estimates of landings and discards. (* Indicates including BMS)

| Year | Landings | Discards |
| :---: | :---: | :---: |
| 1991 | 5656 | 13231 |
| 1992 | 5321 | 11874 |
| 1993 | 4781 | 9854 |
| 1994 | 5732 | 11028 |
| 1995 | 5588 | 9170 |
| 1996 | 7072 | 9356 |
| 1997 | 5167 | 5893 |
| 1998 | 4986 | 10863 |
| 1999 | 5356 | 11065 |
| 2000 | 5445 | 6611 |
| 2001 | 2020 | 1536 |
| 2002 | 3118 | 4154 |
| 2003 | 5968 | 5520 |
| 2004 | 6434 | 883 |
| 2005 | 5239 | 505 |
| 2006 | 2756 | 386 |
| 2007 | 3347 | 2242 |
| 2008 | 4222 | 2104 |
| 2009 | 3241 | 1556 |
| 2010 | 3404 | 907 |
| 2011 | 1860 | 152 |
| 2012 | 686 | 26 |
| 2013 | 889 | 1065 |
| 2014 | 1845 | 332 |
| 2015 | 2510 | 554 |
| 2016 | 2504 | 401 |
| 2017 | 4430 | 379 |
| 2018 | 3850 | 788 |
| 2019 | 7782* | 303 |
| 2020 | 5510* | 130 |
| 2021 | 4095* | 1117 |

Table 7.3.4. Haddock in 6.b. International landings numbers ( ${ }^{*} \mathbf{1 0}^{3}$ ) at-age.

| Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 87 | 6807 | 3011 | 1344 | 558 | 32 | 464 |
| 1992 | 86 | 3642 | 5623 | 964 | 580 | 364 | 160 |
| 1993 | 28 | 1919 | 4740 | 1157 | 489 | 144 | 290 |
| 1994 | 30 | 1160 | 5299 | 3665 | 1039 | 66 | 141 |
| 1995 | 1 | 146 | 5205 | 4791 | 1319 | 279 | 43 |
| 1996 | 2 | 5149 | 1861 | 4149 | 2347 | 473 | 85 |
| 1997 | 0 | 319 | 2102 | 2155 | 3658 | 1540 | 192 |
| 1998 | 4 | 392 | 1815 | 1340 | 1898 | 2284 | 1301 |
| 1999 | 245 | 2600 | 2994 | 1972 | 1228 | 1600 | 2291 |
| 2000 | 33 | 3446 | 5081 | 3006 | 1296 | 1176 | 1963 |
| 2001 | 402 | 994 | 1116 | 555 | 991 | 462 | 549 |
| 2002 | 657 | 2983 | 3998 | 2111 | 809 | 217 | 392 |
| 2003 | 920 | 8103 | 11010 | 1848 | 1189 | 879 | 593 |
| 2004 | 197 | 1765 | 9502 | 9119 | 1364 | 286 | 472 |
| 2005 | 887 | 2835 | 6866 | 7913 | 725 | 98 | 182 |
| 2006 | 2344 | 768 | 1290 | 2356 | 2269 | 428 | 150 |
| 2007 | 31 | 1220 | 2709 | 1074 | 1550 | 1634 | 719 |
| 2008 | 17 | 749 | 6191 | 1164 | 479 | 761 | 848 |
| 2009 | 5 | 11 | 244 | 5243 | 460 | 261 | 486 |
| 2010 | 0 | 71 | 196 | 352 | 4078 | 274 | 294 |
| 2011 | 2 | 23 | 71 | 177 | 181 | 2405 | 222 |
| 2012 | 0 | 0 | 134 | 51 | 0 | 35 | 410 |
| 2013 | 162 | 14 | 2 | 46 | 6 | 46 | 553 |
| 2014 | 226 | 1553 | 418 | 52 | 138 | 47 | 679 |
| 2015 | 9 | 820 | 3214 | 104 | 7 | 61 | 112 |
| 2016 | 127 | 612 | 2137 | 842 | 3 | 2 | 11 |
| 2017 | 7 | 1336 | 1783 | 2179 | 1207 | 58 | 59 |
| 2018 | 0 | 3418 | 502 | 2233 | 598 | 222 | 13 |
| 2019 | 10 | 1514 | 10556 | 59 | 484 | 90 | 60 |
| 2020 | 21 | 1936 | 1190 | 3392 | 364 | 518 | 180 |
| 2021 | 132 | 544 | 2863 | 556 | 1788 | 53 | 337 |

Table 7.3.5. Haddock in 6.b. International discards numbers (* ${ }^{*} 0^{3}$ ) at-age.

| YEAR | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 21099 | 27040 | 12178 | 3998 | 1146 | 313 | 58 |
| 1992 | 15998 | 21069 | 12961 | 4397 | 1181 | 312 | 46 |
| 1993 | 11151 | 17456 | 10755 | 3781 | 1128 | 317 | 69 |
| 1994 | 8140 | 19464 | 12570 | 4545 | 1409 | 410 | 91 |
| 1995* | 2748 | 9685 | 16379 | 4965 | 1145 | 508 | 36 |
| 1996 | 12094 | 13662 | 9051 | 5463 | 952 | 278 | 7 |
| 1997* | 9957 | 10216 | 3286 | 1944 | 1344 | 218 | 15 |
| 1998* | 14220 | 19415 | 8357 | 3423 | 1842 | 483 | 91 |
| 1999* | 17037 | 19348 | 9209 | 3526 | 2191 | 1084 | 485 |
| 2000* | 8189 | 9136 | 5616 | 1912 | 755 | 322 | 103 |
| 2001* | 7268 | 1019 | 583 | 266 | 50 | 15 | 21 |
| 2002 | 12706 | 8136 | 539 | 334 | 89 | 43 | 51 |
| 2003 | 5655 | 15503 | 3558 | 217 | 97 | 48 | 8 |
| 2004 | 735 | 2346 | 781 | 93 | 22 | 10 | 2 |
| 2005 | 174 | 888 | 554 | 210 | 28 | 11 | 11 |
| 2006 | 536 | 707 | 336 | 58 | 22 | 8 | 1 |
| 2007 | 1458 | 8609 | 921 | 440 | 678 | 193 | 0 |
| 2008 | 458 | 1458 | 5246 | 128 | 28 | 203 | 82 |
| 2009 | 218 | 696 | 993 | 2803 | 35 | 2 | 18 |
| 2010* | 152 | 463 | 868 | 1736 | 19 | 2 | 2 |
| 2011* | 2 | 36 | 4 | 6 | 0 | 174 | 27 |
| 2012* | 5 | 6 | 10 | 7 | 3 | 0 | 18 |
| 2013* | 4733 | 84 | 99 | 40 | 33 | 38 | 12 |
| 2014* | 179 | 1454 | 0 | 0 | 0 | 0 | 0 |
| 2015* | 71 | 2153 | 173 | 0 | 0 | 0 | 0 |
| 2016* | 245 | 439 | 503 | 146 | 0 | 0 | 0 |
| 2017* | 1187 | 334 | 20 | 12 | 0 | 0 | 0 |
| 2018* | 88 | 2955 | 3 | 40 | 0 | 0 | 0 |
| 2019* | 275 | 471 | 308 | 8 | 76 | 0 | 0 |
| 2020* | 237 | 263 | 0.00 | 0.2 | 0 | 0 | 0 |
| 2021* | 2797 | 1556 | 339 | 72 | 74 | 0 | 0 |

[^3]Table 7.3.6. Haddock in 6.b. International catch (landings and discards) numbers (* ${ }^{*} \mathbf{0}^{3}$ ) at-age.

| Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 21186 | 33847 | 15189 | 5341 | 1704 | 346 | 522 |
| 1992 | 16084 | 24711 | 18584 | 5361 | 1761 | 676 | 206 |
| 1993 | 11178 | 19375 | 15494 | 4938 | 1617 | 461 | 359 |
| 1994 | 8170 | 20623 | 17868 | 8209 | 2449 | 476 | 232 |
| 1995 | 2749 | 9831 | 21584 | 9756 | 2464 | 787 | 79 |
| 1996 | 12096 | 18811 | 10911 | 9612 | 3299 | 751 | 92 |
| 1997 | 9957 | 10535 | 5388 | 4098 | 5002 | 1758 | 206 |
| 1998 | 14224 | 19807 | 10173 | 4763 | 3740 | 2767 | 1391 |
| 1999 | 17282 | 21949 | 12203 | 5499 | 3419 | 2684 | 2776 |
| 2000 | 8222 | 12581 | 10698 | 4917 | 2050 | 1498 | 2066 |
| 2001 | 7669 | 2013 | 1699 | 821 | 1041 | 477 | 570 |
| 2002 | 13363 | 11119 | 4537 | 2445 | 898 | 260 | 444 |
| 2003 | 6576 | 23606 | 14568 | 2065 | 1286 | 927 | 602 |
| 2004 | 932 | 4112 | 10282 | 9212 | 1386 | 296 | 474 |
| 2005 | 1061 | 3723 | 7420 | 8124 | 753 | 109 | 193 |
| 2006 | 2880 | 1475 | 1626 | 2414 | 2291 | 436 | 151 |
| 2007 | 1489 | 9829 | 3630 | 1514 | 2227 | 1827 | 720 |
| 2008 | 476 | 2207 | 11437 | 1291 | 507 | 964 | 930 |
| 2009 | 223 | 707 | 1237 | 8046 | 495 | 263 | 504 |
| 2010 | 152 | 534 | 1064 | 2087 | 4096 | 276 | 296 |
| 2011 | 4 | 59 | 75 | 183 | 181 | 2579 | 249 |
| 2012 | 5 | 6 | 144 | 58 | 3 | 35 | 428 |
| 2013 | 4896 | 98 | 101 | 86 | 39 | 84 | 565 |
| 2014 | 406 | 3008 | 418 | 52 | 138 | 47 | 679 |
| 2015 | 80 | 2973 | 3387 | 104 | 7 | 61 | 112 |
| 2016 | 374 | 1051 | 2639 | 988 | 3 | 2 | 11 |
| 2017 | 1194 | 1670 | 1802 | 2191 | 1207 | 58 | 59 |
| 2018 | 88 | 6373 | 504 | 2273 | 598 | 222 | 13 |
| 2019 | 288 | 1995 | 10866 | 67 | 560 | 90 | 60 |
| 2020 | 264 | 2202 | 1190 | 3392 | 364 | 518 | 180 |
| 2021 | 2931 | 2101 | 3202 | 628 | 1862 | 53 | 337 |

Table 7.3.7. Haddock in 6.b. International landings mean weights-at-age (kg).

| YEAR | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 0.302 | 0.402 | 0.444 | 0.592 | 0.724 | 0.963 | 0.704 |
| 1992 | 0.136 | 0.366 | 0.455 | 0.658 | 0.612 | 0.759 | 0.954 |
| 1993 | 0.305 | 0.402 | 0.503 | 0.701 | 0.830 | 0.820 | 0.972 |
| 1994 | 0.314 | 0.356 | 0.452 | 0.558 | 0.638 | 1.224 | 0.890 |
| 1995 | 0.377 | 0.311 | 0.414 | 0.479 | 0.640 | 0.699 | 1.236 |
| 1996 | 0.327 | 0.436 | 0.501 | 0.487 | 0.627 | 0.709 | 0.783 |
| 1997 | 0.300 | 0.315 | 0.401 | 0.444 | 0.564 | 0.661 | 0.973 |
| 1998 | 0.256 | 0.344 | 0.494 | 0.517 | 0.542 | 0.591 | 0.678 |
| 1999 | 0.085 | 0.177 | 0.326 | 0.417 | 0.495 | 0.595 | 0.662 |
| 2000 | 0.111 | 0.206 | 0.242 | 0.328 | 0.413 | 0.483 | 0.720 |
| 2001 | 0.094 | 0.281 | 0.344 | 0.497 | 0.427 | 0.522 | 0.690 |
| 2002 | 0.107 | 0.196 | 0.227 | 0.323 | 0.521 | 0.627 | 0.804 |
| 2003 | 0.100 | 0.164 | 0.246 | 0.350 | 0.387 | 0.423 | 0.606 |
| 2004 | 0.142 | 0.172 | 0.241 | 0.293 | 0.446 | 0.617 | 0.754 |
| 2005 | 0.103 | 0.184 | 0.230 | 0.310 | 0.461 | 0.614 | 1.095 |
| 2006 | 0.084 | 0.167 | 0.223 | 0.327 | 0.440 | 0.598 | 0.789 |
| 2007 | 0.096 | 0.238 | 0.275 | 0.322 | 0.449 | 0.521 | 0.578 |
| 2008 | 0.125 | 0.197 | 0.302 | 0.444 | 0.583 | 0.752 | 0.984 |
| 2009 | 0.300 | 0.346 | 0.420 | 0.416 | 0.692 | 0.512 | 1.031 |
| 2010 | 0.052 | 0.420 | 0.517 | 0.457 | 0.591 | 0.980 | 1.473 |
| 2011 | 0.214 | 0.329 | 0.613 | 0.454 | 0.694 | 0.594 | 0.780 |
| 2012 | 0.189 | 0.368 | 0.632 | 0.850 | 0.898 | 1.412 | 1.238 |
| 2013 | 0.510 | 0.554 | 0.713 | 0.972 | 1.361 | 0.948 | 1.267 |
| 2014 | 0.186 | 0.351 | 0.268 | 0.545 | 1.000 | 1.036 | 1.370 |
| 2015 | 0.107 | 0.327 | 0.615 | 0.354 | 1.178 | 0.948 | 1.439 |
| 2016 | 0.409 | 0.574 | 0.664 | 0.767 | 1.576 | 1.808 | 2.650 |
| 2017 | 0.173 | 0.460 | 0.587 | 0.692 | 0.944 | 0.780 | 1.270 |
| 2018 | -1 | 0.332 | 0.564 | 0.705 | 0.935 | 1.235 | 1.928 |
| 2019 | 0.190 | 0.489 | 0.589 | 0.825 | 1.116 | 1.440 | 1.683 |
| 2020 | 0.298 | 0.531 | 0.576 | 0.807 | 0.749 | 1.029 | 1.363 |
| 2021 | 0.284 | 0.394 | 0.512 | 0.837 | 0.818 | 1.138 | 1.150 |

Table 7.3.8. Haddock in 6.b. International discards mean weights-at-age (kg).

| YEAR | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 0.142 | 0.199 | 0.253 | 0.306 | 0.345 | 0.358 | 0.499 |
| 1992 | 0.133 | 0.217 | 0.258 | 0.298 | 0.330 | 0.342 | 0.499 |
| 1993 | 0.137 | 0.220 | 0.260 | 0.307 | 0.346 | 0.359 | 0.504 |
| 1994 | 0.153 | 0.226 | 0.263 | 0.308 | 0.345 | 0.356 | 0.508 |
| 1995 | 0.118 | 0.220 | 0.276 | 0.325 | 0.341 | 0.329 | 0.438 |
| 1996 | 0.136 | 0.218 | 0.276 | 0.326 | 0.370 | 0.348 | 0.515 |
| 1997 | 0.136 | 0.238 | 0.272 | 0.312 | 0.372 | 0.442 | 0.512 |
| 1998 | 0.141 | 0.248 | 0.267 | 0.291 | 0.327 | 0.336 | 0.451 |
| 1999 | 0.139 | 0.212 | 0.255 | 0.288 | 0.313 | 0.318 | 0.417 |
| 2000 | 0.189 | 0.267 | 0.289 | 0.311 | 0.330 | 0.334 | 0.484 |
| 2001 | 0.135 | 0.247 | 0.294 | 0.344 | 0.412 | 0.440 | 0.513 |
| 2002 | 0.137 | 0.254 | 0.308 | 0.335 | 0.398 | 0.338 | 0.382 |
| 2003 | 0.161 | 0.223 | 0.287 | 0.342 | 0.337 | 0.440 | 0.487 |
| 2004 | 0.148 | 0.218 | 0.282 | 0.343 | 0.324 | 0.371 | 0.449 |
| 2005 | 0.171 | 0.240 | 0.298 | 0.357 | 0.387 | 0.473 | 0.511 |
| 2006 | 0.132 | 0.233 | 0.334 | 0.420 | 0.495 | 0.435 | 0.423 |
| 2007 | 0.115 | 0.179 | 0.233 | 0.227 | 0.243 | 0.280 | 0.420 |
| 2008 | 0.202 | 0.264 | 0.279 | 0.370 | 0.351 | 0.358 | 0.446 |
| 2009 | 0.247 | 0.287 | 0.319 | 0.343 | 0.360 | 0.662 | 0.507 |
| 2010 | 0.141 | 0.220 | 0.292 | 0.301 | 0.322 | 0.534 | 0.250 |
| 2011 | 0.178 | 0.248 | 0.300 | 0.302 | 0.795 | 0.727 | 0.481 |
| 2012 | 0.263 | 0.295 | 0.488 | 0.319 | 0.339 | 0.733 | 0.797 |
| 2013 | 0.201 | 0.337 | 0.228 | 0.397 | 0.247 | 0.679 | 0.980 |
| 2014 | 0.082 | 0.218 | - | - | - | - | - |
| 2015 | 0.104 | 0.227 | 0.334 | - | - | - | - |
| 2016 | 0.240 | 0.276 | 0.325 | 0.393 | - | - | - |
| 2017 | - | 0.308 | 0.482 | 0.520 | 0.726 | - | - |
| 2018 | 0.088 | 0.258 | 0.361 | 0.422 | 0.479 | 0.536 | - |
| 2019 | 0.180 | 0.259 | 0.297 | 0.374 | 0.486 | - | - |
| 2020 | 0.2422 | 0.274 | - | 0.512 | - | - | - |
| 2021 | 0.182 | 0.284 | 0.336 | 0.344 | 0.394 | - | - |

Table 7.3.9. Haddock in 6.b. International catch (landings and discards) mean weights-at-age (kg).

| YEAR | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | 0.142 | 0.240 | 0.291 | 0.378 | 0.469 | 0.414 | 0.681 |
| 1992 | 0.133 | 0.239 | 0.318 | 0.362 | 0.423 | 0.567 | 0.852 |
| 1993 | 0.137 | 0.238 | 0.335 | 0.400 | 0.493 | 0.503 | 0.882 |
| 1994 | 0.153 | 0.233 | 0.319 | 0.420 | 0.469 | 0.477 | 0.740 |
| 1995 | 0.118 | 0.222 | 0.309 | 0.401 | 0.501 | 0.460 | 0.870 |
| 1996 | 0.136 | 0.278 | 0.314 | 0.396 | 0.553 | 0.575 | 0.762 |
| 1997 | 0.136 | 0.240 | 0.322 | 0.381 | 0.512 | 0.634 | 0.940 |
| 1998 | 0.141 | 0.250 | 0.308 | 0.354 | 0.436 | 0.546 | 0.663 |
| 1999 | 0.138 | 0.208 | 0.272 | 0.334 | 0.379 | 0.483 | 0.619 |
| 2000 | 0.189 | 0.250 | 0.267 | 0.321 | 0.382 | 0.451 | 0.709 |
| 2001 | 0.133 | 0.264 | 0.326 | 0.447 | 0.427 | 0.520 | 0.683 |
| 2002 | 0.135 | 0.239 | 0.237 | 0.325 | 0.509 | 0.579 | 0.755 |
| 2003 | 0.153 | 0.203 | 0.256 | 0.349 | 0.384 | 0.424 | 0.604 |
| 2004 | 0.147 | 0.198 | 0.244 | 0.294 | 0.444 | 0.609 | 0.753 |
| 2005 | 0.114 | 0.197 | 0.235 | 0.311 | 0.459 | 0.600 | 1.062 |
| 2006 | 0.093 | 0.198 | 0.245 | 0.329 | 0.441 | 0.595 | 0.787 |
| 2007 | 0.114 | 0.186 | 0.265 | 0.294 | 0.386 | 0.496 | 0.578 |
| 2008 | 0.199 | 0.241 | 0.291 | 0.437 | 0.571 | 0.669 | 0.937 |
| 2009 | 0.248 | 0.288 | 0.339 | 0.391 | 0.668 | 0.513 | 1.012 |
| 2010 | 0.141 | 0.247 | 0.333 | 0.327 | 0.590 | 0.977 | 1.464 |
| 2011 | 0.198 | 0.280 | 0.596 | 0.449 | 0.695 | 0.603 | 0.748 |
| 2012 | 0.263 | 0.295 | 0.622 | 0.784 | 0.372 | 1.411 | 1.219 |
| 2013 | 0.211 | 0.368 | 0.236 | 0.704 | 0.423 | 0.827 | 1.261 |
| 2014 | 0.140 | 0.286 | 0.268 | 0.545 | 1.000 | 1.036 | 1.370 |
| 2015 | 0.104 | 0.254 | 0.601 | 0.354 | 1.178 | 0.948 | 1.439 |
| 2016 | 0.298 | 0.449 | 0.600 | 0.711 | 1.556 | 1.808 | 2.650 |
| 2017 | 0.219 | 0.430 | 0.586 | 0.691 | 0.944 | 0.780 | 1.270 |
| 2018 | 0.088 | 0.298 | 0.563 | 0.700 | 0.935 | 1.233 | 1.928 |
| 2019 | 0.180 | 0.434 | 0.581 | 0.771 | 1.030 | 1.440 | 1.683 |
| 2020 | 0.245 | 0.500 | 0.576 | 0.807 | 0.749 | 1.029 | 1.363 |
| 2021 | 0.186 | 0.312 | 0.493 | 0.781 | 0.801 | 1.138 | 1.150 |

Table 7.3.10. Haddock 6.b. Stock mean weights-at-age (kg). (* Indicates values from survey data)

| YEAR | AGE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0* | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1991 | NA | 0.142 | 0.240 | 0.291 | 0.378 | 0.469 | 0.414 | 0.681 |
| 1992 | NA | 0.133 | 0.239 | 0.318 | 0.362 | 0.423 | 0.567 | 0.852 |
| 1993 | NA | 0.137 | 0.238 | 0.335 | 0.400 | 0.493 | 0.503 | 0.882 |
| 1994 | NA | 0.153 | 0.233 | 0.319 | 0.420 | 0.469 | 0.477 | 0.740 |
| 1995 | NA | 0.137 | 0.234 | 0.314 | 0.392 | 0.471 | 0.484 | 0.805 |
| 1996 | NA | 0.136 | 0.242 | 0.319 | 0.396 | 0.488 | 0.516 | 0.821 |
| 1997 | NA | 0.136 | 0.242 | 0.320 | 0.399 | 0.506 | 0.530 | 0.839 |
| 1998 | NA | 0.137 | 0.245 | 0.314 | 0.390 | 0.494 | 0.538 | 0.795 |
| 1999 | 0.017 | 0.134 | 0.240 | 0.305 | 0.373 | 0.476 | 0.540 | 0.771 |
| 2000 | NA | 0.148 | 0.245 | 0.297 | 0.357 | 0.452 | 0.538 | 0.739 |
| 2001 | 0.022 | 0.148 | 0.242 | 0.299 | 0.368 | 0.427 | 0.527 | 0.723 |
| 2002 | 0.028 | 0.147 | 0.242 | 0.282 | 0.356 | 0.426 | 0.516 | 0.686 |
| 2003 | 0.029 | 0.150 | 0.233 | 0.272 | 0.355 | 0.416 | 0.491 | 0.674 |
| 2004 | NA | 0.151 | 0.231 | 0.266 | 0.347 | 0.429 | 0.517 | 0.701 |
| 2005 | 0.028 | 0.136 | 0.220 | 0.260 | 0.345 | 0.444 | 0.546 | 0.771 |
| 2006 | 0.027 | 0.128 | 0.207 | 0.243 | 0.322 | 0.447 | 0.561 | 0.792 |
| 2007 | 0.030 | 0.124 | 0.197 | 0.249 | 0.315 | 0.423 | 0.545 | 0.757 |
| 2008 | 0.013 | 0.134 | 0.204 | 0.256 | 0.333 | 0.460 | 0.594 | 0.823 |
| 2009 | 0.008 | 0.154 | 0.222 | 0.275 | 0.352 | 0.505 | 0.574 | 0.875 |
| 2010 | NA | 0.159 | 0.232 | 0.295 | 0.355 | 0.531 | 0.650 | 0.956 |
| 2011 | 0.008 | 0.180 | 0.248 | 0.365 | 0.380 | 0.582 | 0.651 | 0.948 |
| 2012 | 0.018 | 0.210 | 0.270 | 0.436 | 0.477 | 0.579 | 0.834 | 1.076 |
| 2013 | 0.024 | 0.212 | 0.295 | 0.425 | 0.531 | 0.550 | 0.866 | 1.141 |
| 2014 | 0.020 | 0.191 | 0.295 | 0.411 | 0.562 | 0.616 | 0.971 | 1.212 |
| 2015 | 0.023 | 0.183 | 0.297 | 0.465 | 0.567 | 0.734 | 0.965 | 1.207 |
| 2016 | 0.018 | 0.203 | 0.330 | 0.465 | 0.619 | 0.906 | 1.206 | 1.588 |
| 2017 | 0.012 | 0.195 | 0.357 | 0.458 | 0.601 | 1.020 | 1.080 | 1.598 |
| 2018 | 0.021 | 0.170 | 0.343 | 0.524 | 0.600 | 1.123 | 1.161 | 1.731 |
| 2019 | 0.012 | 0.178 | 0.373 | 0.586 | 0.645 | 1.129 | 1.242 | 1.794 |
| 2020 | 0.015 | 0.206 | 0.422 | 0.581 | 0.736 | 1.043 | 1.258 | 1.779 |
| 2021 | 0.025 | 0.184 | 0.395 | 0.560 | 0.750 | 0.892 | 1.124 | 1.479 |

Table 7.3.11. Haddock in 6.b. Scottish Q3 Rockall haddock survey 1991-2009 (old index).

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 14458 | 16398 | 4431 | 683 | 315 | 228 | 37 | 64 | 3 |
| 1992 | 20336 | 44912 | 14631 | 3150 | 647 | 127 | 200 | 4 | 32 |
| 1993 | 15220 | 37959 | 15689 | 3716 | 1104 | 183 | 38 | 73 | 21 |
| 1994 | 23474 | 13287 | 11399 | 4314 | 969 | 203 | 30 | 12 | 4 |
| 1995 | 16923 | 16971 | 6648 | 5993 | 1935 | 483 | 200 | 16 | -1 |
| 1996 | 33578 | 19420 | 5903 | 1940 | 1317 | 325 | 69 | 6 | 1 |
| 1997 | 28897 | 10693 | 2384 | 538 | 292 | 281 | 71 | 9 | 1 |
| 1998 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1999 | 10178 | 9969 | 2410 | 708 | 279 | 172 | 90 | 64 | 32 |
| 2000 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 2001 | 31813 | 7455 | 521 | 284 | 154 | 39 | 14 | 12 | 14 |
| 2002 | 11704 | 20925 | 2464 | 173 | 105 | 65 | 20 | 10 | 15 |
| 2003 | 2526 | 10114 | 10927 | 1656 | 138 | 97 | 100 | 26 | 6 |
| 2004 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 2005 | 24452 | 4082 | 920 | 1506 | 2107 | 231 | 33 | 13 | 7 |
| 2006 | 3570 | 18715 | 2562 | 256 | 1402 | 1694 | 349 | 16 | 6 |
| 2007 | 558 | 2671 | 6019 | 570 | 254 | 516 | 367 | 28 | 2 |
| 2008 | 85 | 560 | 966 | 3813 | 182 | 41 | 282 | 249 | 49 |
| 2009 | 132 | 139 | 323 | 488 | 1651 | 40 | 9 | 54 | 17 |

Table 7.3.12. Haddock in 6.b. Scottish Q3 Rockall haddock survey 2011 onwards (number per 10 hours).
Mean

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 5.34 | 15.86 | 137.60 | 17.92 | 67.95 | 101.45 | 816.59 | 8.03 |
| 2012 | 14778.60 | 2.15 | 8.47 | 55.82 | 9.59 | 59.30 | 32.03 | 424.06 |
| 2013 | 3247.62 | 12258.74 | 7.94 | 22.05 | 36.56 | 22.59 | 27.98 | 347.17 |
| 2014 | 1925.84 | 6146.09 | 5274.52 | 3.84 | 0.00 | 8.82 | 0.00 | 109.53 |
| 2015 | 1211.67 | 2237.97 | 5390.05 | 4194.88 | 0.00 | 0.00 | 8.60 | 51.27 |
| 2016 | 33441.08 | 1154.50 | 1403.12 | 2444.32 | 1702.92 | 13.55 | 0.76 | 25.63 |
| 2017 | 18583.48 | 23852.74 | 615.22 | 966.59 | 1595.60 | 691.67 | 0.71 | 10.78 |
| 2018 | 6118.72 | 2878.79 | 10395.64 | 249.22 | 532.29 | 856.83 | 325.10 | 3.94 |
| 2019 | 2933.15 | 4003.82 | 2934.86 | 5806.47 | 107.41 | 131.20 | 317.02 | 178.34 |
| 2020 | 25149.28 | 1456.96 | 2114.08 | 774.48 | 1700.02 | 39.63 | 52.60 | 94.52 |
| 2021 | 29363.38 | 9445.42 | 679.64 | 864.30 | 414.42 | 892.77 | 45.34 | 55.80 |

Variance

| $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\mathbf{2 0 1 1} 2.90$ | 27.76 | 1697.44 | 19.08 | 340.97 | 526.65 | 25921.61 | 2.69 |  |
| $\mathbf{2 0 1 2} 108959685.11$ | 0.68 | 5.41 | 273.30 | 6.75 | 277.59 | 44.99 | 6162.86 |  |
| $\mathbf{2 0 1 3}$ | 625196.73 | 3346529.04 | 25.18 | 152.47 | 844.77 | 86.09 | 360.46 | 16760.42 |
| $\mathbf{2 0 1 4} 195960.84$ | 403331.21 | 620653.05 | 5.04 | 0.00 | 7.89 | 0.00 | 462.41 |  |
| $\mathbf{2 0 1 5}$ | 65123.93 | 61454.67 | 171518.52 | 314384.40 | 0.00 | 0.00 | 12.67 | 99.08 |
| 2016 | 549457752.18 | 21367.73 | 29421.70 | 130113.84 | 88936.29 | 6.86 | 0.25 | 34.44 |
| 2017 | 184354785.46 | 4911190.53 | 6874.48 | 14332.22 | 41415.90 | 6184.16 | 0.02 | 3.33 |
| 2018 | 1204812.27 | 116784.67 | 1341249.83 | 5489.44 | 14867.12 | 62230.53 | 14816.24 | 1.29 |
| 2019 | 803305.84 | 260170.82 | 192851.61 | 2520643.85 | 804.91 | 820.30 | 6292.20 | 1548.22 |
| 2020 | 136236520.83 | 33883.47 | 50157.51 | 16396.79 | 131875.36 | 172.91 | 304.24 | 554.79 |
| 2021 | 77305629.48 | 1045147.35 | 5115.86 | 10912.53 | 2708.44 | 12225.15 | 63.14 | 24.66 |

Table 7.3.13. Haddock in Division 27.6.b. Survey biomass index (estimate and variance) including ages 0+in kg per 10 hours.

| Year | idx | var |
| :--- | :--- | :--- |
| 2011 | 667.64 | 11323.48 |
| 2012 | 822.24 | 44562.36 |
| 2013 | 3139.49 | 173139.86 |
| 2014 | 2908.21 | 69487.92 |
| 2015 | 4059.65 | 85357.36 |
| 2016 | 3545.33 | 245237.44 |
| 2017 | 7219.27 | 238626.71 |
| 2018 | 5982.50 | 267038.84 |
| 2019 | 6176.63 | 916841.05 |
| 2020 | 3558.91 | 122493.71 |
| 2021 | 4464.18 | 99515.50 |

Table 7.3.14. Haddock in Division 27.6.b. Annual estimates of modal length, Lc, LF=M and Lmax.

| Year | Lmode | Lc | LFeqM | Lmax |
| :--- | :--- | :--- | :--- | :--- |
| 2012 | 48 | 34 | 39.25 | 74 |
| 2013 | 26 | 24 | 31.75 | 74 |
| 2014 | 27 | 26 | 33.25 | 82 |
| 2015 | 29 | 26 | 33.25 | 73 |
| 2016 | 38 | 31 | 37 | 72 |
| 2017 | 37 | 34 | 39.25 | 76 |
| 2018 | 31 | 28 | 34.75 | 91 |
| 2019 | 37 | 33 | 38.5 | 82 |
| 2020 | 38 | 33 | 38.5 | 82 |
| 2021 | 33 | 30 | 36.25 |  |

Table 7.3.15. Haddock in Division 27.6.b. Summary of the stock indicators used in the assessment and advice.

| Year | low | value | high | Lbar | LBI | landings | discards | bms | catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | NA | NA | NA | NA | NA | 5656 | 13231 | 0 | 18868 |
| 1992 | NA | NA | NA | NA | NA | 5321 | 11874 | 0 | 17199 |
| 1993 | NA | NA | NA | NA | NA | 4781 | 9854 | 0 | 14655 |
| 1994 | NA | NA | NA | NA | NA | 5732 | 11028 | 0 | 16751 |
| 1995 | NA | NA | NA | NA | NA | 5588 | 9170 | 0 | 14754 |
| 1996 | NA | NA | NA | NA | NA | 7072 | 9356 | 0 | 16433 |
| 1997 | NA | NA | NA | NA | NA | 5167 | 5893 | 0 | 11049 |
| 1998 | NA | NA | NA | NA | NA | 4986 | 10863 | 0 | 15841 |
| 1999 | NA | NA | NA | NA | NA | 5356 | 11065 | 0 | 16417 |
| 2000 | NA | NA | NA | NA | NA | 5445 | 6611 | 0 | 12058 |
| 2001 | NA | NA | NA | NA | NA | 2020 | 1536 | 0 | 3554 |
| 2002 | NA | NA | NA | NA | NA | 3118 | 4154 | 0 | 7274 |
| 2003 | NA | NA | NA | NA | NA | 5968 | 5520 | 0 | 11498 |
| 2004 | NA | NA | NA | NA | NA | 6434 | 883 | 0 | 7321 |
| 2005 | NA | NA | NA | NA | NA | 5239 | 505 | 0 | 5740 |
| 2006 | NA | NA | NA | NA | NA | 2756 | 386 | 0 | 3141 |
| 2007 | NA | NA | NA | NA | NA | 3347 | 2242 | 0 | 5587 |
| 2008 | NA | NA | NA | NA | NA | 4222 | 2104 | 0 | 6325 |
| 2009 | NA | NA | NA | NA | NA | 3241 | 1556 | 0 | 4799 |
| 2010 | NA | NA | NA | NA | NA | 3404 | 907 | 0 | 4311 |
| 2011 | 46 | 67 | 88 | NA | NA | 1860 | 152 | 0 | 2012 |
| 2012 | 41 | 82 | 124 | 45.7 | 1.22 | 686 | 26 | 0 | 712 |
| 2013 | 232 | 314 | 396 | 42.9 | 1.15 | 889 | 1065 | 0 | 1952 |
| 2014 | 239 | 291 | 342 | 44.8 | 1.2 | 1845 | 332 | 0 | 2175 |
| 2015 | 349 | 406 | 463 | 39.5 | 1.05 | 2510 | 554 | 0 | 3063 |
| 2016 | 257 | 355 | 452 | 39.1 | 1.04 | 2504 | 401 | 0 | 2906 |
| 2017 | 626 | 722 | 818 | 40.6 | 1.09 | 4430 | 379 | 0 | 4810 |
| 2018 | 497 | 598 | 700 | 38.7 | 1.03 | 3850 | 788 | 0 | 4640 |
| 2019 | 430 | 618 | 805 | 41.4 | 1.1 | 7778 | 303 | 4 | 8090 |
| 2020 | 287 | 356 | 424 | 41.2 | 1.1 | 5508 | 130 | 2 | 5640 |
| 2021 | 385 | 446 | 508 | 39.0 | 1.04 | 4094 | 1117 | 1 | 5212 |



Figure 7.3.1. Haddock in Division 6.b. Official landings by country.


Figure 7.3.2. Haddock in Division 6.b. ICES estimated landings as submitted to Intercatch for 2021.


Figure 7.3.3. Haddock in Division 6.b. ICES estimated landings and discards after raising in Intercatch (grey=imported landings; black=imported discards; red=raised discards).

## Total Catch Numbers At Length



Figure 7.3.4. Haddock in Division 6.b. Catch numbers-at-length by sampled and un-sampled landings and sampled and raised (unsampled) discards for 2021, after allocations within InterCatch.


Figure 7.3.5. Haddock in Division 6.b. ICES estimates of total landings and discards.


Figure 7.3.6. Haddock in Division 6.b. Discard proportion by weight.


Figure 7.3.7. Haddock in Division 6.b. Catch-at-length in numbers by year. Red: discards, blue: landings.


Figure 7.3.8. Haddock in Division 6.b. Mean weight-at-age in catch, discards and landings.


Figure 7.3.9. Haddock in Division 6.b. Mean weight-at-age in catch, stock and survey.


Figure 7.3.10. Haddock in Division 27.6.b. Log mean standardised index values by cohort from ScoRoc-Q3 survey (2011 onwards).

log index

Figure 7.3.11. Haddock in Division 27.6.b.. Within-survey correlations for the ScoRoc-Q3 survey, comparing index values at different ages for the same cohorts (2011 onwards).


Figure 7.3.12. Haddock in Divisino 27.6.b. Comparison of survey index at age between the new index (2011 onwards) and the 'standard index' from the 2021 WG.


Figure 7.3.13. Haddock in Division 27.6.b. Comparison of biomass index with SSB from 2021 WG assessment (including the intermediate year value). Series are mean standardised.


Figure 7.3.14. Haddock in Divison 27.6.b. Survey biomass (kg $10 \mathbf{~ h r}^{-1}$ ) over time by age.


Figure 7.3.15. Haddock in Division 27.6.b. von Bertalanffy curves fitted to survey data by cohort. Note that circles represent individual length-age combinations and the weighting of each value is not shown.


Figure 7.3.16. Comparison of growth parameters estimated at WG 2022 with those in Fishbase (other stocks around UK).


Figure 7.3.17. Haddock in Division 27.6.b. Summary of the stock assessment.

## 8 Haddock in Divisions 7.b,c,e-k

## Type of assessment in 2021

The Celtic Sea haddock ( $27.7 \mathrm{~b}, \mathrm{c}, \mathrm{e}-\mathrm{k}$ ) assessment was benchmarked in 2020, with discard and landings data reviewed and updated from 2005 onwards.

The 2022 SAM assessment was undertaken in the web tool: www.stockassessment.org. The procedure detailed in the Stock Annex, performed in the preceding year was followed.

## ICES advice applicable to 2022

Last year's full advice is available in the Report of the ICES Advisory Committee, 2021. ICES Advice 2021, had.27.7b-k. https://doi.org/10.17895/ices.advice.7764. The headline advice was as follows:
"ICES advises that when the MSY approach is applied, catches in 2022 should be no more than 15946 tonnes. ICES notes the existence of a precautionary management plan, developed and adopted by some of the relevant management authorities for this stock."

### 8.1 General

## Stock description and management units



The basis for the stock assessment area 7.b,c,e-k is described in detail in the stock annex.
Figure 8.1 shows the spatial distribution of international haddock landings in the NE Atlantic for 2016. It is clear from the figure that the stock extends into Area 8 and it could be argued that landings from 8 should be included in the stock area. In recent years these landings varied between 20 and 300 t which is up to $4 \%$ of the total landings in the stock area.

The TAC for haddock is set for the combined Areas $7 . \mathrm{b}-\mathrm{k}, 8,10$ and 10 and EU waters of CECAF 34.1.1. This does not correspond to the stock assessment area (7.b-k).

## 2022 management (Council Regulation (EU) 2022/109)

| Species: | Haddock <br> Melanogrammus aeglefinus | Zone:7b-k, 8, 9 and 10; Union waters of <br> CECAF 34.1.1 <br> (HAD 7 PXA34) |
| :--- | :--- | :--- |
| Belgium | 146 | Analytical TAC <br> France |
| Ireland | 8762 | Article 8(2) of this Regulation applies |
| Union | 2920 |  |
| United Kingdom | 11828 |  |
| TAC | 2550 |  |

Since 2009, a separate TAC is set for 7.a haddock; previously a separate allocation for 7.a existed within the TAC for $7,8,9$ and 10 .

The 2020 EU Council Regulation included Article 13, "Remedial measures for cod and whiting in the Celtic Sea" which will impact the Celtic Sea haddock fishery as these three species occupy similar areas. Article 13 implements spatial and fishing gear restrictions in an effort to reduce fishing pressure on cod and whiting.

### 8.2 The fishery

The official landings reported to ICES are given in Table 8.1. Before 2002, the TAC was well in excess of the landings in the TAC area. The TAC appeared to become restrictive for France in 2003-2004 and Ireland in 2001-2003. During 2005-2008 landings were well below the TAC. In 2009 and 2010, the total landings were still below the TAC, but the quota appeared to become restrictive again for Ireland and Belgium. Since 2011, the TAC has been close to the total landings, and can be assumed to be restrictive for all countries. In the last two years uptake by France has reduced to less than $50 \%$ and UK quota share increased substantially due to Brexit.

Figure 8.1 shows the distribution of international landings between 2015 and 2019. Most haddock landings were taken from the northern North Sea, Irish Sea, Rockall and from the Celtic Sea.

Figure 8.2 shows a longer time-series of official landings and TAC. The time-series is characterised by a number of peaks with rapid increases in the landings, mostly followed by rapid decreases within a few years, suggesting the fishery was taking advantage of sporadic events of very high recruitment. During the 1960s and 1970s, three such peaks in landings occurred: the landings increased from less than 4000 t to 10000 t or more. During the 1980s and early 1990s, landings were relatively stable around 2000-4000 t . During the mid-1990s the haddock landings increased again to over 10000 t , mirroring increased landings in the Irish Sea in that period. Since the late 1990s the landings have varied between 7000 and 10000 t and in 2012, the landings were the highest on record at more than 18000 t .

Working Group estimates of the landings and discards are given in Table 8.2 and shown in Figure 8.3. The discard estimate for 2010 was the highest on record at 16547 tonnes, this was mainly a consequence of the 2009 cohort entering the fishery.
Table 8.3 and Figure 8.4 show that Irish commercial LPUE was relatively low between 2003 and 2007 after which it increased. Effort in the French gadoid fleet has declined considerably since the early 2000s as the result of a decommissioning scheme. The French and Irish 7.fgh fleets both showed an increase in LPUE as the strong 2009 cohort entered the fishery. These data are presented for auxiliary information only; these fleets are not used directly in the assessment.

### 8.2.1 Information from the industry

No updated information from industry was received.

### 8.3 Data

### 8.3.1 Landings and discard numbers-at-age

Catch sampling in 2021 increased compared to 2020 when it was impacted owing to the COVID 19 pandemic and is considered sufficient to describe the stock.

Discard and retained catch-at-age distributions are shown in Figure 8.5. Many of the discarded fish will be above the MLS, which is likely to be the result of restrictive quota

Landings numbers-at-age are given in Table 8.4 and discard numbers-at-age are given in Table 8.5. Despite some uncertainty about the quality of the discard data, it is possible to track strong year classes in both the discards and the landings-at-age matrices. Figure 8.6 shows proportional representation of landings relative to catch (discards + landings) by age, 1993-2021. Discards account for a large proportion of the catch numbers up to age 3 . Figure 8.7 shows the proportions-at-age that are discarded.

Sampled and unsampled catch (landings and discards) by country are shown in Figure 8.8.
Figure 8.9 shows that the raw stock weights-at-age which are fairly noisy, a 3 -year running average was applied to the stock weights used in the assessment. There appear to be cyclical trends in the weights-at-age that follow cohorts (rather than year-effects).

### 8.3.2 Biological

The assumptions of natural mortality and maturity are described in the stock annex. The maturity ogive used in the assessment is quite sharp, with $0.39 \%$ of 2 year olds and $91 \%$ of 3 year olds mature (stock annex).

### 8.3.3 Surveys and commercial tuning fleets

The available surveys and commercial tuning fleets are described in detail in the stock annex. One survey index is used in the assessment: the FR-IRL-IBTS index, which is a combined index from the French EVHOE Q4 WIBTS and Irish IGFS Q4 WIBTS surveys. This is standardised following the VAST procedure (stock annex).

The index data are given Table 8.6. The standardised indices are given by year in Figure 8.10 and by cohort in Figure 8.11 . Figure 8.12 shows the scatterplot matrices of the log indices. These plots indicate that the internal consistency of the indices is robust.

### 8.4 Historical stock development

Model used: SAM
Software used: Stock Assessment.Org (https://www.stockassessment.org)

### 8.4.1 Data screening

The general approach to data screening and analysis was followed in addition to the data exploration tools available in the FLR package FLEDA. The results of the data screening are fully documented using R markdown and are available in the folder 'Data $\backslash$ Stock $\backslash$ had- 7 bce- k ' in the ICES SharePoint.

### 8.4.2 Final update assessment

The final assessment was run with the same settings as established by WKCELTIC 2020 and described in the stock annex. While discards were combined with the landings and not supplied separately to the model, annual discard fractions were incorporated.

Figure 8.13 shows the residuals of that catch proportions-at-age. For age classes where discards dominate, the residuals are relatively large. There are no obvious pattern in the younger ages but the residuals in the middle of the time-series show a mostly positive evolution from the 2006 cohort. The strongest negativities residuals occur for the older age classes in 2006. Observed and assessment predicted catches are shown in Figure 8.14. The predicted catches were generally accurate while there was a tendency for under estimation from 2011-2018. The observed and predicted index CPUE values are shown in Figure 8.15. The assessment generally follows the survey index trends in age classes across the time-series.

In the proportions-at-age residual plots of the survey (FRA-IRL-WIBTS_VAST) there are no consistent patterns (Figure 8.16). The assessment generally follows the survey index trends in age classes across the time-series.

The SAM assessment is shown in Figure 8.17, detailing catch, landings, SSB F and recruits with $95 \%$ confidence intervals.

### 8.4.3 State of the stock

Table 8.7 shows the estimated fishing mortality-at-age and Table 8.8 shows the stock numbers-at-age. The stock summary is given in Table 8.9.

The spawning-stock biomass (SSB) peaked in 2011 as the very strong 2009 year class matured; this cohort was followed by three years of below-average recruitment which led to a rapid decline in SSB after 2011. Recent recruitment has varied around the average, with a notable peak in 2009 and in 2018. SSB appears to have stabilised, while fishing mortality ( F ) has been above Fmsy for the entire time-series but shows a declining trend.

### 8.5 Short-term projections

Because recruitment of haddock is characterised by sporadic events, the assumed median recruitment for the intermediate years introduces significant uncertainty for the SSB estimate.

Short-term projections were performed in SAM as a stochastic process. Recruitment was estimated at 275943 in 2022 and 2023 respectively, (medians 1993-2021; thousands). The short-term predictions are expected to give a reasonably reliable estimate of landings and discards in 2022 (assuming average F 2019-2021 and average discard patterns seen in 2019-2021).

Intermediate year assumptions are given in Table 8.10. The management options are given in Table 8.11.

### 8.6 MSY evaluations and biological reference points

ICES carried out and evaluation of MSY and PA reference points for this stock at WKCELTIC (ICES, 2020). The results are summarized below:

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY <br> approach | MSY $\mathrm{B}_{\text {trigger }}$ | 12822 | $\mathrm{B}_{\mathrm{pa}}$; in tonnes. | ICES (2020a) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.353 | Based on simulation using a segmented regression stockrecruitment relationship (EqSim) | ICES (2020a) |
| Precautionary approach | $\mathrm{B}_{\text {lim }}$ | 9227 | Lowest observed SSB; in tonnes | ICES (2020a) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 12822 | $\mathrm{B}_{\text {lim }}$ combined with the assessment error; $\mathrm{B}_{\text {lim }} \times \exp$ ( $1.645 \times \sigma$ ); $\sigma=0.20$ (default setting); in tonnes | ICES (2020a) |
|  | $\mathrm{F}_{\text {lim }}$ | 1.40 | F with $50 \%$ probability of $\mathrm{SSB}<\mathrm{Bl}_{\text {lim }}$ | ICES (2020a) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.71 | $\mathrm{F}_{\mathrm{po.5}}$; the F that leads to $\mathrm{SSB} \geq \mathrm{B}_{\text {lim }}$ with $95 \%$ probability | ICES (2020a) |
| EU MAP | MAP <br> MSY $B_{\text {trigger }}$ | 12822 | MSY $\mathrm{B}_{\mathrm{pa}}$; in tonnes | $\begin{aligned} & \text { EU (2019), ICES } \\ & \text { (2020a) } \end{aligned}$ |
|  | MAP Blim | 9227 | Lowest observed SSB; in tonnes | $\begin{aligned} & \text { EU (2019), ICES } \\ & \text { (2020a) } \end{aligned}$ |
|  | MAP $\mathrm{F}_{\mathrm{MSY}}$ | 0.353 | $\mathrm{F}_{\text {MSY }}$ | $\begin{aligned} & \text { EU (2019), ICES } \\ & \text { (2020a) } \end{aligned}$ |
|  | MAP range <br> $\mathrm{F}_{\text {lower }}$ | 0.221 | Consistent with ranges resulting in no more than 5\% reduction in long-term yield compared with MSY | $\begin{aligned} & \text { EU (2019), ICES } \\ & \text { (2020a) } \end{aligned}$ |
|  | MAP range $\mathrm{F}_{\text {upper }}$ | 0.521 | Consistent with ranges resulting in no more than 5\% reduction in long-term yield compared with MSY | $\begin{aligned} & \text { EU (2019), ICES } \\ & \text { (2020a) } \end{aligned}$ |

### 8.7 Management plans

The EU multiannual plan (MAP) for the Western Waters (EU, 2019), incorporating the stock haddock 7.b,c,e-k has been agreed. This MAP "establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks", under article 17 states that "It is appropriate to establish the target fishing mortality ( F ) that corresponds to the objective of reaching and maintaining MSY as ranges of values which are consistent with achieving MSY(FmsY). Those ranges, based on best available scientific advice, are necessary in order to provide flexibility to take account of developments in the scientific advice, to contribute to the implementation of the landing obligation and to take into account the characteristics of mixed fisheries."

### 8.8 Uncertainties and bias in assessment and forecast

### 8.8.1 Landings

Sampling levels of the landed catch for recent years are considered to be sufficient to support current assessment approaches, although the assessment is contingent on the accuracy of the
landings statistics. Catch sampling in 2021 increased compared to 2020 when it was impacted owing to the COVID-19 pandemic and is considered sufficient to describe the stock.

Sampling indicated that stock weights-at-age decreased compared to those used for the 2021 assessment. This may have contributed to reduced SBB estimates in the assessment.

### 8.8.2 Discards

Irish discards have been monitored since 1995. The number of trips sampled has varied considerably over time (between three and 62 trips per year). Sample numbers were particularly low in 1995, 1999-2002 and in 2006. During the remaining years, the number of sampled trips was considered sufficient to give reliable estimates of discards.

French discard data exist from 2004 onwards but the data are not considered to be reliable before 2008. The time-series of French discards was reconstructed by assuming that $90 \%$ of one-year olds, $50 \%$ of two-year olds and $10 \%$ of three year olds were discarded throughout the time-series. These proportions were estimated from the available discard and retained catch data provided by France. Discards were estimated for the early part of the time-series at WKROUND (2012) and retained by WKCELTIC up to 2004.

Although recent discard estimates are considered to be more reliable, the problem remains that the number of observer trips is very small compared to the total number of trips (typically $<1 \%$ of all trips are sampled). The level of uncertainty owing to the small sample sizes is likely to be high, but the cost of increasing discard coverage would be considerable. As mentioned sampling levels were considerably low in 2020.

### 8.8.3 Assessment bias

Figure 8.18 shows the retrospective of the ASAP analysis. The predicted catch shows little retrospective pattern neither does the SSB estimate with the Mohn's rho for SSB estimated to be low at $4 \%$. The Recruitment however, has a relatively high Mohn's rho at $21 \%$ owing primarily to the last of five data reductions. F shows variable tendencies with removal of data years, however no overall pattern is discernible and the Mohn's rho is low at $-4 \%$.

The historical assessment results (Figure 8.19) shows a revision in estimated stock size for the 2022 assessment due to the addition of new data for 2021, recent low recruitment and older year classes being removed from the stock.

## $8.9 \quad$ Forecast

The 2018 cohort is projected to account for $32 \%$ the projected catch in 2023, This strong cohort was picked up by both the Irish and French quarter 4 surveys in 2018 but its contribution only accounts for $15 \%$ of SSB in 2024.

Figure 8.20 shows the assessment and forecast of the final SAM run for the FMSY catch option leading to an SSB of 48157 tonnes in 2024 and advised catch of 11901 tonnes.

The assumed recruitment in 2022 and 2023 used in the forecast would constitute a minor part of the projected catches in 2023 ( $8 \%$ ) and approximately $31 \%$ of the SSB in 2024 (Figure 8.21).

### 8.10 Recommendation for next benchmark

### 8.10.1 Stock audit

The audit of the 2021 report did not raise any concerns.

### 8.10.2 Recommendations for future work

Future benchmarks should consider mixed fisheries and multispecies interactions as well as environmental drivers that may be impacting on growth and recruitment of all three species.

Catch data should continue to be monitored for indirect evidence of improved selection patterns due to the augmented TCMs in the Celtic Sea. Direct monitoring of escapement through SMPs would also be useful.

It would be desirable to include discards separately in the assessment model in order to specify greater precision for the discard numbers-at-age than for the landings numbers-at-age. However, WKROUND (2012) concluded that this resulted in undesirable residual patterns. The benchmark workshop did not have sufficient time to fully evaluate this problem.

### 8.11 Management considerations

The stock size fluctuates strongly over the time. The size of the stock is determined to a large extent by recruitment, which has been erratic and in 2018 is shown to have been large. There is no discernible relationship between stock size and recruitment, as is the case with most haddock stocks.

Fishing mortality has been consistently above Fmsy, but this has not led to a decreasing trend in stock size, which suggests that the stock is robust to overfishing, however $F$ has been increasing since 2015 and at current levels the SSB could quickly fall below MSYB ${ }_{\text {trigger }}$ if recruitment were to be low for three or four years. The high recruitment seen in 2018 is moving through the fishery and the older year classes are being removed from the stock.

Discarding of undersize as well as marketable fish is a serious problem for this stock, with approximately $2 / 3$ in catch numbers and almost half the catch weight has been discarded on average over the past decade. Alternative or complimentary approaches to managing such strong, re-cruit-driven fluctuations are required, especially with regard to the EU landings obligation.

The minimum landing size of haddock is 30 cm , which is approximately the same as the mean length of two-year old haddock in the Celtic sea. Because gadoids are caught in a mixed fishery, restrictive quota in recent years have led to increased discarding of marketable fish as well as already considerable discarding of undersized fish. Technical measures have been introduced to reduce discards of undersize gadoids ( 110 mm square-mesh panel in the Nephrops fisheries and 100 mm in the gadoid fisheries). It is not clear whether this is sufficient to reduce discard mortality of future cohorts. It is important that technical measures are fully implemented and their effectiveness in reducing discards and impact on commercial catches are monitored and evaluated.

### 8.12 References

EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing

Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008.

COUNCIL REGULATION (EU) 2020/123 of 27 January 2020, fixing for 2020 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters.

COUNCIL REGULATION (EU) 2021/703 of 26 April 2021, amending Regulations (EU) 2021/91 and (EU) 2021/92 as regards certain fishing opportunities for 2021 in Union and non-Union waters.

ICES. 2016a. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

ICES. 2016b. EU request to ICES to provide Fmsy ranges for selected stocks in ICES subareas 5 to 10. ICES Advice 2016 Book 5, ICES Special Request Advice, Published 5 February 2016.

Table 8.1. Haddock in 7.b,c, e-k. Official landings (quota uptake in brackets).

| Year | BEL | ESP | FRA | IRL | UK* | Others | Total | TAC** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 123 | 0 | 2788 | 908 | 240 | 17 | 4076 |  |
| 1995 | 189 (28\%) | 19 | 2964 (74\%) | 966 (72\%) | 266 (44\%) | 64 | 4468 | 6000 |
| 1996 | 133 (9\%) | 48 | 4527 (49\%) | 1468 (47\%) | 439 (31\%) | 38 | 6653 | 14000 |
| 1997 | 246 (16\%) | 54 | 6581 (71\%) | 2789 (90\%) | 569 (41\%) | 31 | 10270 | 14000 |
| 1998 | 142 (6\%) | 260 | 3674 (28\%) | 2788 (63\%) | 445 (22\%) | 52 | 7361 | 20000 |
| 1999 | 51 (2\%) | 88 | 2725 (19\%) | 2034 (42\%) | 278 (13\%) | 71 | 5247 | 22000 |
| 2000 | 90 (5\%) | 110 | 3088 (28\%) | 3066 (83\%) | 289 (17\%) | 13 | 6656 | 16600 |
| 2001 | 165 (12\%) | 646 | 4842 (61\%) | 3608 (135\%) | 422 (35\%) | 19 | 9702 | 12000 |
| 2002 | 132 (128\%) |  | 4348 (70\%) | 2188 (106\%) | 315 (34\%) | 106 | 7089 | 9300 |
| 2003 | 118 (130\%) |  | 5781 (106\%) | 1867 (103\%) | 393 (48\%) | 82 | 8241 | 8185 |
| 2004 | 136 (127\%) |  | 6130 (96\%) | 1715 (80\%) | 313 (33\%) | 159 | 8453 | 9600 |
| 2005 | 167 (130\%) |  | 4174 (54\%) | 2037 (80\%) | 292 (25\%) | 197 | 6867 | 11520 |
| 2006 | 99 (77\%) |  | 3191 (42\%) | 1874 (73\%) | 274 (24\%) | 183 | 5621 | 11520 |
| 2007 | 119 (93\%) |  | 4143 (54\%) | 1931 (75\%) | 385 (33\%) | 50 | 6628 | 11520 |
| 2008 | 109 (84\%) |  | 3638 (47\%) | 1800 (70\%) | 566 (49\%) | 121 | 6234 | 11579 |
| 2009 | 131 (102\%) |  | 5430 (70\%) | 2983 (116\%) | 716 (62\%) | 48 | 9308 | 11579 |
| 2010 | 170 (132\%) |  | 6240 (81\%) | 2609 (101\%) | 852 (74\%) | 128 | 9999 | 11579 |
| 2011 | 211 (143\%) |  | 8389 (95\%) | 3323 (112\%) | 1657 (124\%) | 129 | 13709 | 13316 |
| 2012 | 232 (125\%) |  | 11793 (106\%) | 4129 (112\%) | 1901 (114\%) | 166 | 18221 | 16645 |
| 2013 | 174 (111\%) |  | 8747 (93\%) | 2699 (86\%) | 1455 (103\%) | 23 | 13098 | 14148 |
| 2014 | 99 (94\%) |  | 6375 (101\%) | 2092 (99\%) | 785 (83\%) | 21 | 9372 | 9479 |
| 2015 | 118 (127\%) |  | 5679 (102\%) | 1657 (89\%) | 769 (92\%) | 6 | 8229 | 8342 |
| 2016 | 88 (109\%) |  | 4487 (93\%) | 1730 (107\%) | 692 (95\%) | 27 | 7024 | 7258 |
| 2017 | 110 (128\%) |  | 4885 (95\%) | 1677 (97\%) | 690 (89\%) | 12 | 7374 | 7751 |
| 2018 | 89 (116\%) |  | 4470 (97\%) | 1444 (94\%) | 583 (84\%) | 9 | 6595 | 6910 |
| 2019 | 90 (97\%) |  | 4259 (77\%) | 1323 (71\%) | 516 (62\%) | 74 | 6262 | 8329 |
| 2020 | 106 (88\%) |  | 3522 (49\%) | 2203 (91\%) | 543 (50\%) | 102 | 6476 | 10859 |
| 2021 | 156 (94\%) |  | 4249 (48\%) | 3379 (114\%) | 515 (21\%) | 149 | 8447 | 15000 |

* UK Includes Channel Islands.
** TAC Applied to subareas 7-10 from 1995 to 2008 and to $7 b-k, 8,9$ and 10 from 2009 onwards.

Table 8.2. Haddock in 7.b,c, e-k. ICES estimate of the landings (lan) and discards (dis).

| Year | $\begin{aligned} & \text { BEL } \\ & \text { Lan } \end{aligned}$ | $\begin{aligned} & \text { ESP } \\ & \text { Lan } \end{aligned}$ | FRA <br> Lan | $\begin{aligned} & \text { IRL } \\ & \text { Lan } \end{aligned}$ | $\begin{aligned} & \text { UK } \\ & \text { Lan } \end{aligned}$ | Others Lan | Total Lan | $\begin{aligned} & \text { FRA } \\ & \text { Dis* } \end{aligned}$ | IRL Dis** | Others Dis*** | Total Dis**** | Total <br> CatCH |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 |  |  |  |  |  |  | 3348 | 505 | 594 | 109 | 1208 | 4556 |
| 1994 |  |  |  |  |  |  | 4131 | 1116 | 594 | 176 | 1886 | 6017 |
| 1995 |  |  |  |  |  |  | 4470 | 730 | 1221 | 267 | 2218 | 6688 |
| 1996 |  |  |  |  |  |  | 6756 | 3170 | 713 | 426 | 4309 | 11065 |
| 1997 |  |  |  |  |  |  | 10827 | 2129 | 502 | 253 | 2883 | 13710 |
| 1998 |  |  |  |  |  |  | 7928 | 680 | 140 | 114 | 934 | 8862 |
| 1999 |  |  |  |  |  |  | 4970 | 477 | 54 | 55 | 586 | 5556 |
| 2000 |  |  |  |  |  |  | 7499 | 1587 | 727 | 189 | 2503 | 10002 |
| 2001 |  |  |  |  |  |  | 9278 | 2234 | 743 | 441 | 3418 | 12696 |
| 2002 | 134 | 85 | 3878 | 2070 | 301 | 20 | 6488 | 871 | 5651 | 552 | 7073 | 13561 |
| 2003 | 116 | 82 | 5960 | 1731 | 362 | 41 | 8292 | 1835 | 6941 | 680 | 9456 | 17748 |
| 2004 | 137 | 143 | 6336 | 1785 | 303 | 73 | 8777 | 1108 | 5156 | 486 | 6750 | 15527 |
| 2005 | 166 | 209 | 4101 | 2078 | 285 | 0 | 6839 | 1564 | 5818 | 2571 | 9953 | 16792 |
| 2006 | 98 | 194 | 3131 | 1899 | 269 | 1 | 5592 | 1313 | 2745 | 1841 | 5899 | 11491 |
| 2007 | 117 | 186 | 4134 | 2139 | 385 | 1 | 6961 | 372 | 2483 | 696 | 3552 | 10513 |
| 2008 | 108 | 166 | 4577 | 1984 | 558 | 0 | 7392 | 990 | 3741 | 2930 | 7660 | 15052 |
| 2009 | 129 | 49 | 5503 | 3270 | 711 | 2 | 9664 | 905 | 3320 | 3098 | 7322 | 16986 |
| 2010 | 170 | 115 | 6421 | 2899 | 821 | 3 | 10429 | 3260 | 4570 | 10870 | 18701 | 29130 |
| 2011 | 211 | 78 | 8381 | 3702 | 1551 | 35 | 13957 | 3963 | 4329 | 7515 | 15807 | 29764 |
| 2012 | 232 | 79 | 12293 | 4596 | 1929 | 67 | 19196 | 2754 | 2653 | 2878 | 8285 | 27481 |
| 2013 | 174 | 51 | 8738 | 3097 | 1458 | 20 | 13538 | 671 | 1116 | 2175 | 3962 | 17501 |
| 2014 | 99 | 3 | 6350 | 2543 | 849 | 2 | 9846 | 1732 | 1171 | 2715 | 5619 | 15464 |
| 2015 | 118 | 0 | 5683 | 2035 | 766 | 6 | 8608 | 2024 | 2519 | 2398 | 6941 | 15549 |
| 2016 | 88 | 0 | 4573 | 2271 | 689 | 27 | 7648 | 5482 | 2810 | 3773 | 12065 | 19713 |
| 2017 | 111 | 0 | 4895 | 2381 | 699 | 11 | 8099 | 2633 | 1928 | 2130 | 6691 | 14789 |
| 2018 | 89 | 0 | 4377 | 1989 | 578 | 12 | 7046 | 1920 | 1189 | 2688 | 5798 | 12844 |
| 2019 | 89 | 89 | 4548 | 2412 | 518 | 27 | 7683 | 1616 | 1445 | 542 | 3603 | 11259 |
| 2020 | 102 | 176 | 3815 | 3193 | 546 | 27 | 7859 | 1450 | 1873 | 937 | 4260 | 12119 |
| 2021 | 149 | 108 | 4257 | 4211 | 516 | 19 | 9260 | 706 | 1075 | 604 | 2385 | 11645 |

* For 1993-2007 fixed discard ratios were used to estimate French discards.
** For 1993-1994, the mean Irish discards over 1995-1999 were used.
*** Estimated from the proportion of the landings of `Others' between 1993 and 2012.
**** Discard estimates are available from 2005; prior to 2005, discard estimates are based on limited sampling.

Table 8.3. Haddock in 7.b,c, e-k. LPUE (kg/hour fishing) of haddock and effort (hours fishing $\times$ 1000) for Irish Otter trawls in 7.bc, 7.fgh and 7.jk, the French demersal fleet in 7.bc-ek and effort only for the UK trawl fleets (excluding beam trawls) in 7.e-k (effort in fishing days).

|  | FR GAD <br> 7ek effort | FR GAD <br> 7ek Ipue | IRL OTB <br> 7bc effort | IRL OTB <br> 7bc Ipue | IRL OTB <br> 7fgh effort | IRL OTB <br> 7fgh Ipue | IRL OTB <br> 7jk effort | IRL OTB <br> 7jk Ipue | UK Trawl 7e-k effort |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | NA | NA | NA | NA | NA | NA | NA | NA | 51.5 |
| 1984 | NA | NA | NA | NA | NA | NA | NA | NA | 161.8 |
| 1985 | NA | NA | NA | NA | NA | NA | NA | NA | 143.7 |
| 1986 | NA | NA | NA | NA | NA | NA | NA | NA | 123.5 |
| 1987 | NA | NA | NA | NA | NA | NA | NA | NA | 108.9 |
| 1988 | NA | NA | NA | NA | NA | NA | NA | NA | 112.9 |
| 1989 | NA | NA | NA | NA | NA | NA | NA | NA | 119.9 |
| 1990 | NA | NA | NA | NA | NA | NA | NA | NA | 133.2 |
| 1991 | NA | NA | NA | NA | NA | NA | NA | NA | 118.8 |
| 1992 | NA | NA | NA | NA | NA | NA | NA | NA | 129.9 |
| 1993 | NA | NA | NA | NA | NA | NA | NA | NA | 101.1 |
| 1994 | NA | NA | NA | NA | NA | NA | NA | NA | 88.5 |
| 1995 | NA | NA | 78 | 5.77 | 64 | 1.48 | 106 | 2.20 | 88.1 |
| 1996 | NA | NA | 47 | 4.16 | 60 | 5.35 | 73 | 3.24 | 89.5 |
| 1997 | NA | NA | 63 | 4.36 | 65 | 5.83 | 92 | 8.23 | 101.8 |
| 1998 | NA | NA | 79 | 5.71 | 72 | 4.09 | 99 | 5.88 | 94.6 |
| 1999 | NA | NA | 77 | 5.27 | 51 | 2.35 | 52 | 3.53 | 132.8 |
| 2000 | 306 | 6.12 | 74 | 4.73 | 61 | 10.43 | 72 | 4.25 | 141.1 |
| 2001 | 333 | 10.57 | 78 | 4.30 | 69 | 8.69 | 81 | 7.41 | 117.5 |
| 2002 | 289 | 10.63 | 63 | 2.81 | 79 | 3.22 | 108 | 5.50 | 113.1 |
| 2003 | 264 | 15.15 | 81 | 2.09 | 87 | 3.26 | 123 | 3.88 | 102.4 |
| 2004 | 217 | 19.39 | 82 | 2.51 | 97 | 3.49 | 108 | 3.35 | 105.5 |
| 2005 | 175 | 14.67 | 69 | 2.45 | 127 | 4.53 | 93 | 3.70 | 100.9 |
| 2006 | 167 | 10.64 | 60 | 2.56 | 119 | 4.19 | 89 | 3.59 | 106.3 |
| 2007 | 160 | 14.97 | 60 | 3.31 | 136 | 4.01 | 103 | 3.66 | 113.6 |
| 2008 | 148 | 19.60 | 48 | 4.36 | 127 | 4.56 | 84 | 4.60 | 93.7 |
| 2009 | 150 | 22.65 | 48 | 5.47 | 141 | 9.25 | 82 | 7.09 | 98.6 |


|  | FR GAD <br> 7ek effort | FR GAD <br> 7ek lpue | IRL OTB <br> 7bc effort | IRL OTB <br> 7bc lpue | IRL OTB <br> 7fgh effort | IRL OTB <br> 7fgh lpue | IRL OTB <br> 7jk effort | IRL OTB <br> 7jk lpue | UK Trawl <br> 7e-k effort |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2010 | 131 | 30.83 | 54 | 4.36 | 144 | 7.33 | 101 | 5.15 | 103.7 |
| 2011 | 216 | 22.90 | 40 | 6.39 | 129 | 10.51 | 84 | 5.58 | 87.1 |
| 2012 | 188 | 45.03 | 44 | 4.93 | 135 | 13.17 | 84 | 6.58 | 86.2 |
| 2013 | 215 | 27.40 | 42 | 5.38 | 126 | 8.69 | 80 | 4.92 | 40.3 |
| 2014 | 203 | 19.81 | 46 | 5.22 | 142 | 5.11 | 77 | 3.91 | 32.1 |
| 2015 | NA | NA | 31 | 4.42 | 150 | 4.95 | 78 | 2.91 | 21.2 |
| 2016 | NA | NA | 39 | 2.41 | 164 | 4.94 | 83 | 3.09 | NA |
| 2017 | NA | NA | 36 | 2.25 | 151 | 5.10 | 92 | 2.43 | NA |
| 2018 | NA | NA | 46 | 2.19 | 125 | 5.33 | 93 | 1.70 | NA |
| 2019 | NA | NA | 32 | 2.42 | 127 | 5.86 | 93 | 1.73 | NA |
| 2020 | NA | NA | 34 | 2.80 | 98 | 11.2 | 84 | 1.86 | NA |
|  | NA | 39 | 4.23 | 92 | 14.68 | 86 | 2.70 | NA |  |

Table 8.4. Haddock in 7.b,c, e-k. Landings numbers-at-age.

|  | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 491 | 3291 | 948 | 810 | 255 | 129 | 129 | 45 |
| 1994 | 0 | 1277 | 5223 | 674 | 302 | 94 | 24 | 35 | 16 |
| 1995 | 0 | 4275 | 1622 | 1327 | 270 | 245 | 46 | 0 | 0 |
| 1996 | 0 | 3693 | 15998 | 818 | 313 | 93 | 32 | 10 | 9 |
| 1997 | 0 | 1353 | 9645 | 5553 | 716 | 354 | 139 | 144 | 110 |
| 1998 | 0 | 167 | 3184 | 7403 | 1443 | 307 | 178 | 86 | 61 |
| 1999 | 0 | 476 | 654 | 1464 | 2425 | 307 | 18 | 19 | 6 |
| 2000 | 0 | 2197 | 2996 | 784 | 741 | 1250 | 205 | 35 | 28 |
| 2001 | 0 | 4297 | 8638 | 1131 | 303 | 317 | 321 | 54 | 39 |
| 2002 | 0 | 879 | 4274 | 3400 | 765 | 39 | 89 | 74 | 26 |
| 2003 | 0 | 703 | 8791 | 2160 | 1226 | 116 | 43 | 49 | 51 |
| 2004 | 0 | 125 | 5948 | 4663 | 928 | 589 | 51 | 12 | 20 |
| 2005 | 0 | 1075 | 1732 | 4230 | 1821 | 280 | 75 | 1 | 3 |
| 2006 | 0 | 839 | 3250 | 1034 | 2189 | 484 | 42 | 28 | 0 |
| 2007 | 0 | 404 | 4617 | 2916 | 737 | 1310 | 161 | 33 | 4 |
| 2008 | 0 | 1692 | 3268 | 3736 | 1046 | 286 | 414 | 91 | 50 |
| 2009 | 0 | 338 | 7111 | 2760 | 1890 | 577 | 228 | 234 | 38 |
| 2010 | 0 | 1757 | 5192 | 6031 | 1036 | 580 | 257 | 110 | 123 |
| 2011 | 0 | 100 | 12726 | 3607 | 3410 | 661 | 261 | 129 | 132 |
| 2012 | 0 | 82 | 1135 | 19931 | 2559 | 1795 | 323 | 109 | 108 |
| 2013 | 0 | 86 | 465 | 1899 | 10533 | 861 | 468 | 96 | 44 |
| 2014 | 0 | 277 | 854 | 467 | 1511 | 5585 | 368 | 219 | 40 |
| 2015 | 0 | 41 | 4881 | 632 | 309 | 928 | 2030 | 257 | 80 |
| 2016 | 0 | 62 | 310 | 5200 | 216 | 143 | 546 | 682 | 92 |
| 2017 | 0 | 58 | 2019 | 1071 | 3930 | 135 | 117 | 246 | 312 |
| 2018 | 0 | 70 | 714 | 2833 | 926 | 1653 | 42 | 64 | 150 |
| 2019 | 0 | 513 | 1566 | 1257 | 2678 | 529 | 762 | 41 | 110 |
| 2020 | 0 | 120 | 4318 | 1449 | 755 | 1381 | 260 | 175 | 30 |
| 2021 | 0 | 285 | 1295 | 6691 | 740 | 569 | 640 | 248 | 169 |

Table 8.5. Haddock in 7.b,c, e-k. Discard numbers-at-age.

|  | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 7617 | 2816 | 160 | 6 | 0 | 0 | 0 | 0 |
| 1994 | 0 | 15120 | 3069 | 170 | 5 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 32830 | 1977 | 91 | 4 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 20734 | 8976 | 187 | 9 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 12613 | 10022 | 493 | 5 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 3580 | 2348 | 445 | 5 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 3742 | 1562 | 100 | 10 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 29015 | 2521 | 64 | 3 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 25234 | 6772 | 219 | 2 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 21624 | 20729 | 249 | 7 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 52412 | 11075 | 352 | 8 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 11733 | 21598 | 1395 | 61 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 30472 | 25291 | 6821 | 97 | 1 | 0 | 0 | 0 |
| 2006 | 0 | 20089 | 4529 | 11 | 10 | 4 | 1 | 0 | 0 |
| 2007 | 0 | 10748 | 8498 | 572 | 6 | 6 | 0 | 0 | 0 |
| 2008 | 0 | 34221 | 12620 | 1676 | 78 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 21175 | 13989 | 592 | 64 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 95699 | 19014 | 2742 | 34 | 1 | 0 | 0 | 0 |
| 2011 | 0 | 5881 | 58967 | 1675 | 262 | 16 | 1 | 0 | 1 |
| 2012 | 0 | 2732 | 5169 | 18518 | 153 | 55 | 2 | 0 | 0 |
| 2013 | 0 | 4076 | 2767 | 1372 | 4028 | 58 | 2 | 1 | 1 |
| 2014 | 0 | 20197 | 3315 | 507 | 631 | 732 | 4 | 1 | 0 |
| 2015 | 0 | 3590 | 18090 | 704 | 26 | 155 | 162 | 13 | 6 |
| 2016 | 0 | 27587 | 5222 | 8406 | 51 | 12 | 56 | 501 | 2 |
| 2017 | 0 | 3208 | 11913 | 1602 | 2121 | 31 | 2 | 4 | 3 |
| 2018 | 0 | 5287 | 5127 | 5306 | 491 | 215 | 0 | 2 | 2 |
| 2019 | 0 | 12878 | 2847 | 773 | 409 | 37 | 17 | 1 | 4 |
| 2020 | 0 | 2722 | 10938 | 597 | 28 | 25 | 1 | 1 | 0 |
| 2021 | 0 | 4890 | 3773 | 2799 | 23 | 12 | 1 | 0 | 0 |

Table 8.6. Haddock in 7.b,c, e-k. VAST survey data.

| Year \Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 34982.4 | 194259.7 | 15511.0 | 1334.3 | 1035.4 | 27.7 | 16.2 | 8.8 |
| 2004 | 103867.4 | 19061.2 | 23731.4 | 2359.3 | 957.7 | 523.2 | 886.2 | 10.5 |
| 2005 | 55665.8 | 31406.5 | 4458.3 | 6394.8 | 821.6 | 233.3 | 46.9 | 0.0 |
| 2006 | 31208.7 | 10366.1 | 6855.0 | 1490.4 | 1348.0 | 280.7 | 58.1 | 36.0 |
| 2007 | 247100.9 | 14940.9 | 3707.3 | 2046.6 | 679.5 | 886.7 | 100.0 | 15.6 |
| 2008 | 86672.2 | 55580.3 | 2482.9 | 657.1 | 744.0 | 288.1 | 749.1 | 203.1 |
| 2009 | 877972.9 | 20715.2 | 16571.4 | 592.8 | 357.3 | 310.6 | 403.0 | 185.2 |
| 2010 | 32993.8 | 304206.9 | 10352.3 | 5037.2 | 272.1 | 259.2 | 349.2 | 122.1 |
| 2011 | 20579.7 | 12717.4 | 79367.2 | 2428.1 | 1343.6 | 256.1 | 147.0 | 58.1 |
| 2012 | 7210.7 | 6947.1 | 4289.0 | 14181.3 | 768.1 | 722.3 | 111.8 | 58.9 |
| 2013 | 224645.3 | 2602.7 | 2864.8 | 1441.9 | 5204.1 | 408.2 | 395.8 | 52.3 |
| 2014 | 29933.8 | 57670.6 | 1177.0 | 963.1 | 1019.8 | 2106.0 | 338.0 | 139.1 |
| 2015 | 124666.7 | 27660.7 | 17862.8 | 641.1 | 402.0 | 756.3 | 1232.6 | 88.4 |
| 2016 | 17973.7 | 50953.4 | 13233.5 | 5759.2 | 457.2 | 235.0 | 931.0 | 287.3 |
| 2017 | 49415.3 | 6918.5 | 16135.7 | 3316.2 | 944.4 | 100.0 | 16.6 | 212.8 |
| 2018 | 268416.0 | 9928.8 | 1646.1 | 2772.4 | 1484.4 | 756.6 | 27.5 | 30.3 |
| 2019 | 86436.1 | 144323.1 | 4827.1 | 999.3 | 1753.2 | 561.8 | 342.3 | 26.5 |
| 2020 | 32867.8 | 34934.2 | 54667.9 | 990.0 | 552.9 | 1167.1 | 1263.0 | 375.1 |
| 2021 | 74261.7 | 15950.7 | 14723.3 | 12309.2 | 279.2 | 88.3 | 336.4 | 175.2 |

Table 8.7. Haddock in 7.b,c, e-k. Fishing mortality- (F) at-age.

|  | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | - | 0.336 | 0.733 | 0.58 | 0.577 | 0.568 | 0.549 | 0.611 | 0.611 |
| 1994 | - | 0.326 | 0.704 | 0.552 | 0.542 | 0.529 | 0.511 | 0.569 | 0.569 |
| 1995 | - | 0.322 | 0.698 | 0.552 | 0.539 | 0.525 | 0.508 | 0.563 | 0.563 |
| 1996 | - | 0.312 | 0.687 | 0.554 | 0.545 | 0.529 | 0.512 | 0.566 | 0.566 |
| 1997 | - | 0.324 | 0.725 | 0.611 | 0.622 | 0.614 | 0.602 | 0.662 | 0.662 |
| 1998 | - | 0.317 | 0.718 | 0.616 | 0.648 | 0.652 | 0.647 | 0.705 | 0.705 |
| 1999 | - | 0.298 | 0.681 | 0.583 | 0.618 | 0.626 | 0.625 | 0.674 | 0.674 |
| 2000 | - | 0.326 | 0.761 | 0.656 | 0.703 | 0.721 | 0.722 | 0.762 | 0.762 |
| 2001 | - | 0.332 | 0.791 | 0.692 | 0.753 | 0.775 | 0.783 | 0.818 | 0.818 |
| 2002 | - | 0.32 | 0.78 | 0.684 | 0.758 | 0.789 | 0.805 | 0.841 | 0.841 |
| 2003 | - | 0.308 | 0.754 | 0.673 | 0.758 | 0.84 | 0.873 | 0.911 | 0.911 |
| 2004 | - | 0.31 | 0.758 | 0.673 | 0.748 | 0.838 | 0.869 | 0.886 | 0.886 |
| 2005 | - | 0.301 | 0.717 | 0.605 | 0.632 | 0.673 | 0.667 | 0.671 | 0.671 |
| 2006 | - | 0.257 | 0.599 | 0.495 | 0.501 | 0.529 | 0.523 | 0.553 | 0.553 |
| 2007 | - | 0.242 | 0.572 | 0.478 | 0.464 | 0.476 | 0.466 | 0.5 | 0.5 |
| 2008 | - | 0.243 | 0.591 | 0.513 | 0.494 | 0.499 | 0.494 | 0.549 | 0.549 |
| 2009 | - | 0.224 | 0.556 | 0.508 | 0.505 | 0.517 | 0.514 | 0.581 | 0.581 |
| 2010 | - | 0.204 | 0.517 | 0.494 | 0.506 | 0.536 | 0.544 | 0.632 | 0.632 |
| 2011 | - | 0.184 | 0.476 | 0.481 | 0.514 | 0.566 | 0.591 | 0.71 | 0.71 |
| 2012 | - | 0.174 | 0.451 | 0.474 | 0.521 | 0.59 | 0.628 | 0.776 | 0.776 |
| 2013 | - | 0.164 | 0.424 | 0.442 | 0.485 | 0.553 | 0.594 | 0.754 | 0.754 |
| 2014 | - | 0.151 | 0.399 | 0.422 | 0.454 | 0.523 | 0.565 | 0.736 | 0.736 |
| 2015 | - | 0.138 | 0.369 | 0.406 | 0.436 | 0.501 | 0.548 | 0.731 | 0.731 |
| 2016 | - | 0.137 | 0.365 | 0.414 | 0.448 | 0.512 | 0.553 | 0.744 | 0.744 |
| 2017 | - | 0.133 | 0.364 | 0.424 | 0.471 | 0.539 | 0.568 | 0.759 | 0.759 |
| 2018 | - | 0.127 | 0.353 | 0.418 | 0.463 | 0.531 | 0.55 | 0.739 | 0.739 |
| 2019 | - | 0.11 | 0.312 | 0.387 | 0.44 | 0.512 | 0.534 | 0.724 | 0.724 |
| 2020 | - | 0.096 | 0.272 | 0.349 | 0.409 | 0.483 | 0.495 | 0.67 | 0.67 |
| 2021 |  | 0.098 | 0.277 | 0.361 | 0.437 | 0.531 | 0.544 | 0.735 | 0.735 |

Table 8.8. Haddock in 7.b,c, e-k. Stock numbers-at-age (start of year) (`1000).

|  | Age0 | Age1 | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 137524 | 49768 | 13058 | 4520 | 1235 | 370 | 302 | 189 | 88 |
| 1994 | 392005 | 45842 | 17372 | 3505 | 1575 | 443 | 138 | 116 | 101 |
| 1995 | 475242 | 132728 | 15835 | 4817 | 1253 | 599 | 173 | 55 | 85 |
| 1996 | 165752 | 161429 | 46200 | 4372 | 1722 | 475 | 238 | 70 | 55 |
| 1997 | 59509 | 55715 | 59026 | 12579 | 1551 | 652 | 189 | 98 | 50 |
| 1998 | 87806 | 19777 | 19339 | 16615 | 4021 | 533 | 236 | 70 | 53 |
| 1999 | 359811 | 29275 | 6949 | 5319 | 5684 | 1297 | 182 | 83 | 41 |
| 2000 | 348755 | 122416 | 10656 | 1974 | 1874 | 2006 | 454 | 66 | 44 |
| 2001 | 475463 | 116929 | 42918 | 2772 | 632 | 612 | 643 | 146 | 36 |
| 2002 | 976184 | 159321 | 40207 | 10986 | 845 | 190 | 190 | 195 | 54 |
| 2003 | 241122 | 333972 | 57296 | 9927 | 3503 | 240 | 60 | 58 | 74 |
| 2004 | 341271 | 82024 | 116486 | 15061 | 3092 | 1045 | 66 | 17 | 35 |
| 2005 | 236237 | 113484 | 30131 | 30417 | 4683 | 902 | 274 | 16 | 14 |
| 2006 | 195043 | 79149 | 39074 | 8111 | 9976 | 1558 | 291 | 89 | 11 |
| 2007 | 681770 | 66138 | 29855 | 11916 | 3145 | 3913 | 600 | 117 | 37 |
| 2008 | 414184 | 226343 | 25472 | 9523 | 4513 | 1312 | 1610 | 253 | 66 |
| 2009 | 2316405 | 137399 | 85206 | 7960 | 3596 | 1777 | 556 | 665 | 124 |
| 2010 | 214532 | 782004 | 54019 | 27400 | 3056 | 1421 | 713 | 230 | 306 |
| 2011 | 86560 | 72897 | 305166 | 17851 | 10229 | 1263 | 559 | 280 | 203 |
| 2012 | 60869 | 28896 | 31230 | 104263 | 6815 | 3922 | 494 | 208 | 165 |
| 2013 | 604932 | 20796 | 11901 | 12313 | 39220 | 2628 | 1419 | 178 | 117 |
| 2014 | 225966 | 203091 | 8773 | 4354 | 5532 | 15239 | 1027 | 522 | 95 |
| 2015 | 496568 | 77258 | 84473 | 3331 | 1754 | 2505 | 5935 | 401 | 201 |
| 2016 | 102096 | 169264 | 34136 | 32327 | 1339 | 732 | 1074 | 2286 | 198 |
| 2017 | 143397 | 34294 | 69828 | 13695 | 12646 | 529 | 292 | 424 | 797 |
| 2018 | 869056 | 46502 | 14779 | 26598 | 5712 | 4964 | 199 | 113 | 397 |
| 2019 | 275943 | 294093 | 19140 | 5927 | 10804 | 2355 | 1913 | 78 | 172 |
| 2020 | 186296 | 90507 | 128613 | 7439 | 2491 | 4492 | 964 | 741 | 83 |
| 2021 | 304566 | 63104 | 40065 | 54649 | 3094 | 1043 | 1824 | 404 | 291 |

Table 8.9. Haddock in 7.b,c,e-k. Stock Summary: Estimated recruitment, spawning-stock biomass (SSB), and average fishing mortality.

| Year | R(age 0) | Low | High | SSB | Low | High | $\begin{aligned} & F_{\text {bar }} \\ & (3-5) \end{aligned}$ | Low | High | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 137524 | 67900 | 278540 | 9267 | 6188 | 13880 | 0.575 | 0.394 | 0.84 | 19783 | 13990 | 27975 |
| 1994 | 392005 | 247144 | 621773 | 10974 | 7617 | 15810 | 0.541 | 0.389 | 0.753 | 32099 | 23884 | 43139 |
| 1995 | 475242 | 300708 | 751076 | 11602 | 8259 | 16297 | 0.539 | 0.398 | 0.729 | 46548 | 35747 | 60612 |
| 1996 | 165752 | 106170 | 258771 | 20604 | 15368 | 27625 | 0.543 | 0.407 | 0.724 | 45167 | 36043 | 56599 |
| 1997 | 59509 | 38128 | 92881 | 25319 | 19321 | 33177 | 0.616 | 0.48 | 0.79 | 34879 | 27737 | 43860 |
| 1998 | 87806 | 56018 | 137632 | 19822 | 15556 | 25258 | 0.639 | 0.507 | 0.804 | 25763 | 20960 | 31666 |
| 1999 | 359811 | 231157 | 560067 | 13115 | 10534 | 16328 | 0.609 | 0.485 | 0.764 | 26487 | 21450 | 32706 |
| 2000 | 348755 | 225057 | 540440 | 11742 | 9688 | 14231 | 0.693 | 0.569 | 0.844 | 34376 | 27579 | 42848 |
| 2001 | 475463 | 312325 | 723813 | 18392 | 14106 | 23980 | 0.74 | 0.609 | 0.899 | 41599 | 33481 | 51686 |
| 2002 | 976184 | 650852 | 1464136 | 23438 | 18383 | 29883 | 0.743 | 0.612 | 0.903 | 62197 | 50211 | 77045 |
| 2003 | 241122 | 166214 | 349789 | 27763 | 22230 | 34672 | 0.757 | 0.617 | 0.928 | 73203 | 58226 | 92031 |
| 2004 | 341271 | 235849 | 493815 | 39966 | 31562 | 50608 | 0.753 | 0.599 | 0.947 | 63761 | 52722 | 77111 |
| 2005 | 236237 | 162874 | 342644 | 28620 | 23356 | 35071 | 0.637 | 0.515 | 0.787 | 55332 | 46724 | 65526 |
| 2006 | 195043 | 133225 | 285545 | 24306 | 20233 | 29198 | 0.508 | 0.397 | 0.65 | 45631 | 38857 | 53587 |
| 2007 | 681770 | 472854 | 982989 | 22460 | 19091 | 26424 | 0.473 | 0.372 | 0.601 | 66103 | 53945 | 81000 |
| 2008 | 414184 | 288369 | 594892 | 21044 | 17755 | 24942 | 0.502 | 0.408 | 0.619 | 75801 | 62525 | 91895 |
| 2009 | 2316405 | 1600195 | 3353174 | 32887 | 26829 | 40312 | 0.51 | 0.416 | 0.626 | 190901 | 145238 | 250919 |
| 2010 | 214532 | 138162 | 333116 | 39668 | 33456 | 47033 | 0.512 | 0.418 | 0.626 | 180396 | 141033 | 230746 |
| 2011 | 86560 | 60654 | 123532 | 98217 | 77185 | 124979 | 0.52 | 0.425 | 0.637 | 123094 | 99343 | 152522 |
| 2012 | 60869 | 41822 | 88590 | 71729 | 57593 | 89335 | 0.528 | 0.429 | 0.652 | 82248 | 67253 | 100586 |
| 2013 | 604932 | 421428 | 868340 | 46457 | 37571 | 57445 | 0.493 | 0.402 | 0.606 | 87387 | 72061 | 105973 |
| 2014 | 225966 | 154582 | 330313 | 29415 | 24229 | 35711 | 0.466 | 0.38 | 0.573 | 76165 | 63717 | 91046 |
| 2015 | 496568 | 346622 | 711378 | 40176 | 32827 | 49170 | 0.448 | 0.363 | 0.552 | 94519 | 78256 | 114162 |
| 2016 | 102096 | 70213 | 148455 | 37846 | 31556 | 45390 | 0.458 | 0.371 | 0.566 | 76953 | 64649 | 91599 |
| 2017 | 143397 | 97262 | 211414 | 44690 | 37221 | 53657 | 0.478 | 0.383 | 0.597 | 64019 | 54522 | 75172 |
| 2018 | 869056 | 576472 | 1310138 | 35005 | 29269 | 41866 | 0.471 | 0.374 | 0.593 | 95050 | 75318 | 119951 |
| 2019 | 275943 | 176134 | 432310 | 31040 | 26143 | 36854 | 0.446 | 0.346 | 0.575 | 107357 | 83826 | 137492 |
| 2020 | 186296 | 108999 | 318408 | 56954 | 43126 | 75217 | 0.414 | 0.302 | 0.568 | 86278 | 67596 | 110123 |
| 2021 | 304566 | 132001 | 702725 | 54513 | 40077 | 74148 | 0.443 | 0.305 | 0.643 | 82443 | 61512 | 110497 |

Table 8.10. Haddock in divisions 7.b,c,e-k. Assumptions made for the interim year and in the forecast.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Fages $3-5$ (2022) | 0.443 | Average F = (2019-2021) scaled to Fages 3-5 in 2021 |
| SSB (2023) | 47157 | Short-term forecast; in tonnes |
| $R_{\text {age }} 0(2022,2023)$ | 275943 | Median resampled (1993-2021); in thousands* |
| Total catch (2022) | 15320 | Short-term forecast; in tonnes |
| Projected landings (2022) | 12308 | Short-term forecast, assuming average 2019-2021 landing pattern; in <br> tonnes |
| Projected discards (2022) | 3012 | Short-term forecast, assuming average 2019-2021 discard pattern; in tonnes |

* Random resampling of a distribution may lead to different median estimates.

Table 8.11. Haddock in divisions 7.b,c,e-k. Assumptions made for the interim year and in the forecast.

Haddock in divisions 7.b-k. Annual catch scenarios. All weights are in tonnes.

| Basis | Total catch (2023) | Projected landings (2023) | Projected discards (2023) | $F_{\text {total }}$ <br> (2023) | $F_{\text {projected }}$ landings <br> (2023) | $F_{\text {projected }}$ discards (2023) | $\begin{aligned} & \text { SSB } \\ & (2024) \end{aligned}$ | $\begin{aligned} & \text { \% SSB } \\ & \text { change * } \end{aligned}$ | \% advice change ${ }^{\wedge}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { EU MAP ^^: } \\ & \text { F }_{\text {MSY }} \end{aligned}$ | 11901 | 9064 | 2837 | 0.353 | 0.309 | 0.044 | 48157 | 2.12 | -25 |
| $F=M A P$ | 7862 | 6030 | 1832 | 0.221 | 0.194 | 0.027 | 52430 | 11.2 | -26 |
| $\mathrm{F}_{\text {MSY lower }}$ |  |  |  |  |  |  |  |  |  |
| $F=M A P$ | 16424 | 12419 | 4005 | 0.521 | 0.457 | 0.064 | 43270 | -8.2 | -25.3 |
| $\mathrm{F}_{\text {MSY upper }}$ |  |  |  |  |  |  |  |  |  |
| Other scenarios |  |  |  |  |  |  |  |  |  |
| $F=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 61031 | 29.4 | -100 |
| $\mathrm{F}_{\mathrm{pa}}$ | 20787 | 15604 | 5183 | 0.71 | 0.62 | 0.088 | 38652 | -18.0 | 30 |
| $\mathrm{F}_{\text {lim }}$ | 32583 | 23767 | 8816 | 1.400 | 1.23 | 0.17 | 26386 | -44.0 | 104 |
| $\mathrm{SSB}_{2024}=\mathrm{B}_{\text {lim }}$ | 50807 | 34676 | 16131 | 4.05 | 3.55 | 0.50 | 9227 | -80.4 | 219 |
| $\begin{aligned} & \mathrm{SSB}_{2024}=\mathrm{B}_{\mathrm{pa}}= \\ & \mathrm{MSY}_{\mathrm{trrigger}} \end{aligned}$ | 46662 | 32469 | 14193 | 3.08 | 2.70 | 0.38 | 12822 | -72.8 | 193 |
| $F=F_{2022}$ | 14401 | 10923 | 3478 | 0.44 | 0.39 | 0.06 | 45431 | -3.66 | -9.69 |
| $\begin{aligned} & \mathrm{SSB}_{2024}= \\ & \mathrm{SSB}_{2023} \end{aligned}$ | 12788 | 9731 | 3057 | 0.384 | 0.337 | 0.047 | 47157 | 0.00 | -19.8 |

** Numbers presented are estimations of the reference values.
${ }^{\wedge}$ Advice values for 2022 relative to the corresponding 2021 values (MAP advice of 15 946, 10 570, and 21988 tonnes, respectively; other values are relative to $\mathrm{F}_{\mathrm{ms}}$ ).
$\wedge \wedge$ EU multiannual plan (MAP) for the Western Waters (EU, 2019).


Figure 8.1. International haddock landings by ICES rectangle (all gears; 2016; data from https://stecf.jrc.ec.eu-ropa.eu/data-dissemination).


Figure 8.2. Haddock in 7.b,c,e-k. Official ICES landings and TAC of haddock in 7.b-k.


Figure 8.3. Haddock in 7.b,c,e-k. ICES estimates of landings and quota by country.


Figure 8.4. Haddock in 7.b,c,e-k. Effort ('1000h) of the Irish Otter trawl fleets, the French demersal otter trawl fleet and for UK trawl fleet (effort in fishing days, rescaled to other fleets) and LPUE (kg/h) for the Irish and French fleets.


Figure 8.5. Haddock in 7.b,c,e-k. Discarding by number by age class (grey = landings, white = discards).


Figure 8.6. Haddock in 7.b,c,e-k. Proportional representation of landings relative to catch (discards + landings) by age, 1993-2021.


Figure 8.7. Haddock in 7.b,c,e-k. Proportion of discards by age (left) and year (right).


Figure 8.8. Haddock in 7.b,c,e-k. Distribution sampled and unsampled the catches by country and gear.


Figure 8.9. Haddock in 7.b,c,e-k. Raw stock weights-at-age (left) and the three-year running average stock weights (right).


Figure 8.10. Haddock in 7.b,c,e-k. Log VAST standardised tuning fleets by year. The FRA-IRL-IBTS survey is the combined French EVHOE Q4 WIBTS and Irish IGFS Q4 WIBTS survey.


Figure 8.11. Haddock in 7.b,c,e-k. Log VAST standardised tuning fleets by cohort.

FRA_IRL_WIBTS_VAST

log index

Figure 8.12. Haddock in 7.b,c,e-k. Scatterplot matrix of log indices of cohorts at different ages.


Figure 8.13. Haddock in 7.b,c,e-k. Residuals of the proportions-at-age in catch (upper) and survey (lower).


Figure 8.14. Haddock in 7.b,c,e-k. Observed (line) and predicted (x) catches.


Figure 8.15. Haddock in 7.b,c,e-k. Observed and predicted (circles and line respectively) catch-at-age.




Figure 8.16. Haddock in 7.b,c,e-k. Observed and predicted (circles and line respectively) VAST survey indices.


Figure 8.17. Haddock in 7.b,c,e-k. SAM assessment stock summary plots.


Figure 8.18. Haddock in 7.b,c,e-k. Retrospective analysis of the final SAM assessment run. Catch (top left), SSB (top right), recruitment (bottom left) and $F$ (bottom right).

## SSB (1000 t)



F (ages 3-5)


Rec (age 0; Billions)


Figure 8.19. Haddock 7.b,c,e-k. Historical assessment results (final-year recruitment and SSB assumptions included). The assessment was benchmarked in 2020, prior to which a different method (ASAP based) was applied.


Figure 8.20. Haddock in 7.b,c,e-k. Assessment and forecast of the final SAM run. SSB (top), and F (middle) and recruitment (bottom).


Figure 8.21. Haddock 7.b,c,e-k. Recruitment Contribution of recent year classes used in predictions, and the relative (\%) contributions to catch and SSB (by weight) of these year classes.

## 9 Haddock (Melanogrammus aeglefinus) in Division7.a (Irish Sea)

## Type of assessment

Age-structured assessment model using Age Structured Assessment Program (ASAP).

## ICES advice applicable to 2022

ICES advises that when the MSY approach is applied, catches in 2022 should be no more than 3038 tonnes.

## ICES advice applicable to 2023

ICES advises that when the MSY approach is applied, catches in 2023 should be no more than 2648 tonnes.

### 9.1 General

## Stock descriptions and management units

The stock and management units are both ICES Division 7.a (Irish Sea). Landing taken or reported by Irish vessels in the southern most rectangles of 7. a have been reassigned to the $7 . b-\mathrm{k}$ stock since 2003 because they are believed to be part of the Celtic Sea stock.

## Management applicable to 2023

Management measures include TAC and effort restrictions as well as technical measures. Due to the bycatch of cod in the haddock fishery, the regulations affecting Irish Sea haddock remain linked to those implemented under the cod recovery plan. From 1st January 2019 all fleets catching haddock are subject to the landing obligation.

TAC regulations for 2022 are given below.

2022 management (Council Regulation (EU) 2020/123)

| Species: | Haddock <br> Melanogrammus aeglefinus | Zone: <br> Belgium | (HAD/07A.) |
| :--- | :--- | :--- | :--- |
| France | 43 | Analytical TAC |  |
| Ireland | 196 |  |  |
| Union | 1171 |  |  |
| United Kingdom | 1410 |  |  |
| TAC | 1628 |  |  |

The minimum landing size for haddock in the Irish Sea is 30 cm .

## Landings obligation

Since 2017 the landings obligation has been applied to the stock. According to the delegate regulation (EC, 2015) vessels where more than $25 \%$ of their landings using trawls and seines in the reference years (2013 and 2014) and area were specified gadoids (cod, haddock, whiting and saithe) were covered by the Landings Obligation. This implies that all catches of haddock in the Irish Sea by those vessels must be landed. From the 1st January 2019 all fleets catching haddock are subject to the landings obligation.

## Fishery in 2021

The characteristics of the fishery are described in the stock annex.
The fishery in 2021 was prosecuted by a similar fleet and gears as in recent years, with directed fishing restricted during the cod closure under special conditions. The targeted whitefish fishery that developed during the 1990 using semi-pelagic trawls has declined considerablybut since 2014 there has been a slight increase in activity due to abundance of the haddock stock and increased fishing opportunity. However, this continues to be pursued by a small number of vessel $(<15)$. A proportion of the TAC is taken as bycatch in the Nephrops fishery in a mixed fishery.

In 2020, the whitefish fishery was considerably impacted by the COVID-19 pandemic, resulting in lower fishing effort, which is represented in the landings and total catches.

In 2021, the uptake of TAC was $62 \%$. The primary two nations exploiting the stock are the UK and Ireland. The UK used $54 \%$ of quota allocation whilst Ireland used $99 \%$. ICES catch estimates are adjusted for reallocation of Irish landings from southern rectangles of 7.a to 7.g, as it is believed that these fish do not belong to the 7.a stock. Table 9.1 gives nominal landings of haddock from the Irish Sea (Division 7.a) as reported by each country to ICES since 1984. Newly introduced gear restriction in the Republic of Ireland waters meant that Northern Irish vessels were unable to fish in ROI waters without modifying their gear accordingly.

### 9.2 Data

Sampling was reduced in 2021 due to the COVID-19 pandemic. In the first quarter the TR1 fleet was asked to bring the full final haul ashore and the full haul was sampled following on-sea protocols once the vessel had returned to shore. Sampling on the Nephrops fleet was low during quarter 1, however resumed in quarters $2-4$. The criteria for submitting samples to InterCatch was a minimum of one sample for every $4 \%$ of the landings. If that criterium was not met, sampled data were not submitted. As a result, landings only files were submitted to InterCatch for cod and haddock from Ireland.

In Northern Ireland landings and discards sampling in the first quarter was conducted by requesting the TR1 fleet to bring the full final haul ashore. This was sampled following on-sea protocols once the vessel had returned to shore. Sampling on the Nephrops fleet was low during quarter 1, however resumed in quarters 2-4.

## Landings

Table 9.2 gives the long-term trend of nominal landings of haddock from the Irish Sea (Division 7.a) as reported to ICES since 1972, together with Working Group estimates. The 1993-2005 WG estimates includes sampled-based re-estimates of landings into the main Irish Sea ports. Sampled based evidence suggests that WG estimates are similar to reported landings since 2006. Following the benchmark (WKROUND 2013) the landings have been revised since 1993, and exclude landings from the southern rectangles in the Irish Sea as they not are believed to be part of this stock.

The methods for estimating quantities and composition of haddock landings from 7.a, used in previous years, are described in the stock annex (see Annex 2). The series of numbers-at-age in the international commercial catch is given in Table 9.3. Sampling levels were not considered adequate to derive catch age compositions in 2003.

## Discards

Annual discard data were updated for Ireland and Northern Ireland. Historic discard numbers-at-age for the different sampled fleets are given in the stock annex (see Annex 2). Issues relating to the reliability and confidence in the data were addressed at the benchmark assessment for this stock (WKROUND 2013; WKIrish3 2017).

Methods for estimating quantities and composition of discards from UK (NI) and Irish Nephrops trawlers are described in the stock annex. Sampling levels have increased in recent years. The large estimates of discarding for Nephrops fleets observed by previous WG are still evident. A historic time-series of discard numbers-at-age was constructed at the benchmark. Discard rates are very variable between fleets.

## Biological data

The derivation of biological parameters and variables is described in the stock annex (see Annex 2). Natural mortality-at-age was calculated using the methods proposed by Lorenzen (1996) at WKIrish2 (2016). The proportions mature-at-age was also recalculated at the benchmark, and based on the mean proportion observed during the NIGFS-WIBTS-Q1 survey with a smoother fitted that is updated annually.

There is evidence of trends in mean length-at-age over time (Figure 9.1), which needs to be reflected in the stock weights-at-age. Since 2001 the WG calculated stock weights by fitting a von

Bertalanffy growth curve to survey estimates of mean length-at-age in March, described in the Stock Annex. The procedure was updated this year using NIGFS-WIBTS-Q1 (2021) and quarter one commercial landings data for 2021. The time-series of length-weight parameters indicate a reduction in expected weight-at-length since 1996, although this strength of this decline has reduced in recent years (see stock annex for historical data):

|  | Length-weight parameters |  | Expected weight-at-length |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | A | B | 30 cm | 40 cm |
| 2006 | 0.00506 | 3.165 | 239 | 595 |
| 2007 | 0.00469 | 3.194 | 244 | 612 |
| 2008 | 0.00523 | 3.159 | 242 | 601 |
| 2009 | 0.00431 | 3.224 | 249 | 629 |
| 2010 | 0.00413 | 3.238 | 250 | 635 |
| 2011 | 0.00457 | 3.207 | 250 | 629 |
| 2012 | 0.00499 | 3.174 | 243 | 606 |
| 2013 | 0.00451 | 3.208 | 247 | 622 |
| 2014 | 0.00591 | 3.121 | 241 | 591 |
| 2015 | 0.00423 | 3.232 | 251 | 637 |
| 2016 | 0.00420 | 3.233 | 250 | 634 |
| 2017 | 0.004144 | 3.235 | 249 | 631 |
| 2018 | 0.006453 | 3.108 | 251 | 614 |
| 2019 | 0.004911 | 3.196 | 258 | 647 |
| 2020 | 0.005161 | 3.165 | 245 | 608 |
| 2021 | 0.00591 | 3.1184 | 239 | 586 |

The following parameter estimates were obtained:
Mean $\mathrm{LI}_{\mathrm{yc}}=45.4 \mathrm{~cm} ; \mathrm{K}=0.428 ; \mathrm{t}_{0}=-0.092$

Year-class effects giving estimates of asymptotic length relative to the mean were as follows:

| Year class | Effect | Year class | Effect |
| :---: | :---: | :---: | :---: |
| 1990 | 0.949 | 2004 | 0.983 |
| 1991 | 0.979 | 2005 | 0.989 |
| 1992 | 0.954 | 2006 | 0.953 |
| 1993 | 1.045 | 2007 | 0.986 |
| 1994 | 1.092 | 2008 | 0.961 |
| 1995 | 1.018 | 2009 | 1.002 |
| 1996 | 1.049 | 2010 | 1.058 |
| 1997 | 0.968 | 2011 | 1.074 |
| 1998 | 1.024 | 2012 | 1.106 |
| 1999 | 1.004 | 2013 | 1.014 |
| 2000 | 0.995 | 2014 | 1.019 |
| 2001 | 0.971 | 2015 | 0.943 |
| 2002 | 0.971 | 2016 | 0.920 |
| 2003 | 0.998 | 2017 | 1.001 |
|  |  | 2018 | 0.999 |
|  |  | 2019 | 0.999 |
|  | 2020 |  |  |
| 2021 |  |  |  |

The year-class effects show a smooth decline from the mid-1990s coinciding with the rapid growth of the stock and may represent density-dependent growth effects, although other environmental factors may contribute. There is evidence in a reversal of this trend in recent years. The resultant stock weights-at-age are given in Table 9.3. The weight-at-age in the stock shows a decreasing trend over time which appears to have reversed in recent years.

## Surveys

The survey data considered in the assessment for this stock are given in Table 9.5. All survey series data for haddock available to the Working Group are described in the stock annex (see Annex 2). The following age-structured abundance indices were used in the assessment:

- UK (NI) groundfish survey (NIGFS) in March (age classes 1 to 4, years 1992-2021). Acronym NIGFS-WIBTS-Q1.
- UK (NI) groundfish survey (NIGFS) in October (age classes 0 to 3; years 1991 to 2021). Acronym NIGFS-WIBTS-Q4.
- UK (NI) Methot-Isaacs-Kidd (NI-MIK) net survey in June (age 0; years 1994-2021, excluding 2020).
- UK Fishery Science Partnership (UKFspW) western Irish Sea roundfish survey (age classes 2 to 5, years 2004-2021, the survey was not conducted in 2014).

The relative log standardised indices for cohorts are plotted against time in Figure 9.2. While ages 2 to 4 appear to show strong signal in the UKFspW, the ability to detect the year class in age 5 haddock is less clear. The strong 2013 year class could be tracked in all indices, indicating that the different surveys are capturing the prominent year-class signals in this stock (Figure 9.2). Correlation between survey indices by age is positive for all surveys and show high consistency within each survey (Figure 9.3). The indices from the UKFspW survey in the western Irish Sea also show similar year-class signals to the other survey-series, but are noisy with strong year effects (Figure 9.2).

### 9.3 Assessment

The assessment presented is the single fleet ASAP model.
The following model settings were applied in 2022.

ASAP was used for the assessment and model settings:

| Option | Setting |
| :---: | :---: |
| Use likelihood constant | Yes |
| Mean $F\left(F_{\text {bar }}\right)$ age range | 2-4 |
| Fleet selectivity block 1 | Asymptotic |
| Fleet selectivity block 2 | Age coefficients (age 0-5) (0.2;0.5;0.8;1;0.7;0.5) |
| Fleet selectivity block 3 | Age coefficients (age 0-5) (0.3;0.6;0.7;0.7;0.4;0.2) |
| Fleet selectivity block 4 | Age coefficients (age 0-5) (0.1;0.6;0.8;0.9;1.0;1.0) |
| Discards | Included in catch (not specified separately from landings) |
| Index units | 4 (numbers) |
| Index month | NIGFS-Q1 (3); NIGFS-Q4 (10); NIMIK (7); UKFSPW(3) |
| Index selectivity linked to fleet | -1 (not linked) |
| Index age range | NIGFS-Q1 (1-4); NIGFS-Q4 (0-3); NIMIK (0); UKFSPW(2-5) |
| Index Selectivity (NIGFS-Q1) | Double logistic |
| Index Selectivity (NIGFS-Q4) | Asymptotic |
| Index Selectivity (NIMIK) |  |
| Index Selectivity (UK-FSPW) | Asymptotic |
| Index CV \& ESS (NIGFS-Q1) | Observed strata CV (lower limit 0.1); ESS = 50 |
| Index CV \& ESS (NIGFS-Q4) | Observed strata CV (lower limit 0.1); ESS = 50 |
| Index CV \& ESS (NIMIK) | Observed station CV (lower limit 0.1); ESS = 50; not used for 2020 |
| Index CV \& ESS (UK-FSPW) | $C V=0.7 ; E S S=10$ |
| Phase for F-Mult in 1st year | 1 |
| Phase for F-Mult deviations | 2 |
| Phase for recruitment deviations | 3 |
| Phase for N in 1st Year | 1 |
| Phase for catchability in 1st Year | 3 |
| Phase for catchability deviations | -5 (Assume constant catchability in indices) |
| Phase for unexploited stock size | 1 |
| Phase for steepness | -5 (Do not fit stock-recruitment curve) |
| Catch total CV | $\begin{aligned} & 1993-2000(0.175) ; 2003-2006(0.2) ; 2007-2019 \text { (0.15); } 2020 \text { (0.175); } \\ & 2021 \text { (0.15) } \end{aligned}$ |


| Option | Setting |
| :--- | :--- |
| Catch effective sample size | $1993-2000(50) ; 2003-2006(1) ; 2007-2019(50) ; 2020(1) ; 2021(50)$ |
| Lambda for recruit deviations | 0 (freely estimated) |
| Lambda for total catch | 1 |
| Lambda for F-Mult in 1st year | NA (discards included in catch) |
| Lambda for F-Mult deviations | 0 (freely estimated) |
| Lambda for index | 0 (freely estimated) |
| Lambda for index catchability | 0 for both indices in the model indices (freely estimated) |
| Lambda for catchability devs | NA (phase is negative) |
| Lambda N in 1st year deviations | 0 (freely estimated) |
| Lambda devs initial steepness | 0 (freely estimated) |

## Final update assessment

The final assessment was run with the same settings as established by WKIrish 2017 and described in the stock annex, with the addition of a new selectivity pattern 2013-2021, as applied in 2018 and with the lower starting value for selection of age 0 haddock in the final selectivity block. Hence the changes as described in the stock annex were followed. Discards were combined with the landings as catch in the model.

Figure 9.5 shows the predicted and observed catch. The catch information from 2007 to present is regarded as the most confident, during 2003-2006 it is regarded that catch and sampling information is of relatively lower quality due to lack of sampling opportunity. Before 2003, the catch series is regarded as of intermediate confidence. The model has close fit to the current observed catch 2011-present. Before this time, there is consistent over estimation of the catch 2000-2011 following a period of consistent underestimation of catch 1993-2001. Figure 9.6 shows the residuals of the catch proportions-at-age. For all ages there appears to good fit with no consistent pattern, however, there are some large deviations from observed and predicted for age 5 fish since 2015. Figure 9.7 shows that the catch is dominated by fish $<4$ years, therefore the large residuals for fish of age 5 are likely to result from low sampling and small contribution of $5+$ fish to the stock. The fishing pressure ( F )-at-age is shown in Table 9.6.

The residuals of the indices are shown in Figure 9.7. A good fit to the NI-MIK index is seen across the series, although some single year events are observed with a strong deviation in the last two years of the index. For the UKFSPW survey a poor fit in years 2017 and 2018 is evident. This suggests an inability of the model to track the large survey index values, this should be investigated further to explore the method of index calculation. There is strong tracking of both NIGFS-WIBTS-Q1 and NIGFS-WIBTS-Q4 index patterns in general, however, a general trend to under estimate the NIGFS-WIBTS-Q1 index by the model early 2000s to 2013, followed by a period of over-estimation (during years of high abundance, and with the decline in SSB the model is once again underestimating Q1 survey index.

Figure 9.9 shows the residuals of the survey proportions-at-age. For all indices there is close fit between the observed and model predicted fit for fish up to four years old. The largest deviations occur in five year old fish in the UKFSPW survey, which over-reported five year old fish prior to 2014.

Figure 9.10 shows the retrospective analysis. The predicted catch shows no obvious retrospective pattern, neither does the recruitment estimate or fishing pressure. The results of the assessment are given in Table 9.8.

## Comparison with previous assessments

Figure 9.11 shows the comparison of the current assessment with previous ASAP and model. There is close agreement with the stock trends of the current assessment and the previous assessment. Mohn's Rho values were calculated for five retrospective runs 2021: 2016 for Fbar (0.08), SSB (-0.04) and recruitment (-.51).

## State of the stock

Following a period of sustained decline, since 2008, SSB increased during 2010-2013. A shortterm decline was observed in 2014, but was reversed, and since 2014 the SSB has increased markedly. The stock is characterized by highly variable recruitment. The model indicates above average recruitment for the 2009-2011 year class after below average recruitment for the 2007 and 2008 year classes. Recruitment in 2013 is amongst the highest observed in the time-series and has been followed by strong recruitment in 2014 and 2015. Since 2018 SSB has declined from the highest observed level and continued the decline in 2021.

With the very low recruitment in 2020, the SSB is further projected to decline in 2022 and 2023.

### 9.4 Short-term projections

Short-term projections were performed using FLR libraries. Recruitment for 2022-2024 was estimated at (GM 1993-2019; 364084 thousands). The F used in the forecast for 2022 was derived as $\mathrm{F}_{\mathrm{sq}}=\mathrm{F}_{\text {average }}$ (2018-2021), excluding the 2020 F as this is deemed to be non-representative for the fishery and due to low effort of the TR1 fleet during the COVID-19 pandemic.

Catches were split into landings and discards using the proportions of the catch that were discarded over the full the last three years. Input data for the short-term forecast are given in Table 9.7. The management options output is given in Table 9.9.

Estimates of the relative contribution of recent year classes to the 2023 landings and 2024 SSB are shown in Figure 9.12. The contribution to landings in 2023 consists mainly of the 2018 cohort (72\%), with the SSB in 2024 largely be dependent on the 2019 cohort, comprising $67 \%$ of the SSB and the 2021 cohort contributing $17 \%$. This is an issue as the SSB will largely consist of the plus group.

### 9.5 Biological reference points

## MSY evaluations

In response to an EU special request to provide plausible and updated Fmsy ranges for Irish Sea haddock the management reference points for the stock were re-estimated (Table 9.10 ICES, 2018). The Blim was set as the lowest SBB at which above recruitment in the upper quartile has
been observed (2994t). The S-R plot for Irish Sea haddock shows no obvious S-R relationship mainly because the recruitment is highly variable. Blim was estimated as 4160 t . MSY $B_{\text {trigger }}$ is set to 4281 t as the stock has been fished at or below Fmsy for more than five years. FmSY median point estimates is 0.28 . The upper bound of the FMSY range giving at least $95 \%$ of the maximum yield was estimated to 0.35 and the lower bound at 0.20 . Flim is estimated to be 0.50 as F with $50 \%$ probability of $\mathrm{SSB}<\mathrm{Blim}_{\mathrm{lim}} ; \mathrm{F}_{\mathrm{pa}}$ as $0.41=\mathrm{F}_{\mathrm{p} .05}$ the F that leads to $\mathrm{SSB}>\mathrm{Blim}_{\lim }$ with $95 \%$ probability; $\mathrm{F}_{\text {lim }} \mathrm{x} \exp (-$ $1.645 \times \sigma$ ); $\sigma=0.20$.

## Yield and biomass-per-recruit

Not available for this stock, previous explorations are detailed in the stock annex.

### 9.6 Management plans

There is no specific management plan for haddock in the Irish Sea. The regulations affecting Irish Sea haddock remain linked to those implemented under the cod management plan due to potential for bycatch of cod in a fishery targeting haddock (Council Regulation (EC) 1342/2008).

### 9.7 Uncertainties and bias in assessment and forecast

## Landings

Sampling levels of the landed catch for recent years are considered to be sufficient to support current assessment. However, within the assessment there is relocation of reported landings in rectangles 33 E 2 and 33 E 3 which are not considered part of the stock. Historic misreporting estimates are considered in the assessment and accounted for, current misreporting is not considered to be a factor within the fishery.

## Discards

Sampling levels of discarding at sea remains high. For Northern Irish vessels targeting haddock $27.0 \%$ of trips are observed and $2.7 \%$ of the main Nephrops targeted fishery trips observed, however due to the COVID-19 pandemic the sampling level of the Nephrops targeting fishery has been impacted in 2021.

## Selectivity

A breakpoint in selectivity is applied in 2000, associated with management measures to reduce fishing mortality on cod. The model included three selectivity blocks in fishery-dependent data, reflecting bycatch and targeted fishery until the year 2000 (asymptotic). After 2007, a fleet selectivity pattern without targeted fishing of older fish (dome-shaped) is applied. During 2000-2007 a transition between a fully selected stock to a regime without targeted fishing of older fish is fitted. The use of current specified selectivity blocks may require review at annual at regular intervals. In the current assessment a new selectivity pattern for the fishery was added from 2013 onwards with full selection of fish older than three years. With advice and management for haddock or other species, it is possible that the character of the fishery may change. A retrospective analysis demonstrated a consistent historic downward revision of the perceived SSB trend, however, there is consistent estimation of F. The initial two years of the retrospective plot show significant deviations. This was considered due to the model having a selectivity block, beginning
in 2007, with reduced selection for older fish and the introduction of the UKFspW, with an asymptotic selectivity pattern, starting in 2007. The short period to estimate the selectivity parameters for both the fishery and survey index are considered to contribute to the instability of the model during this time.

## Surveys

The survey indices used in the model have spatial coverage of the assessment area. The combination of a recruitment index (NI-MIK), juvenile fish survey indices (NIGFS-WIBTS-Q1 \& NIGFS-WIBTS-Q4) and the UKFspW survey aimed at older fish using commercial fishing gear means that the full age range of the stock is covered by survey information.

### 9.8 Recommendations for next benchmark assessment

This stock was benchmarked through the WKIrish process in 2016-2017. New estimation of the MikNet survey and re-estimation of ages might need an inter-benchmark.

### 9.9 References

EC. 2015. Commission Delegated Regulation (EU) 2015/2438 of 12 October 2015 establishing a discard plan for certain demersal fisheries in north-western waters.

Table 9.1. Landings ( t ) of HADDOCK in Division 7.a, 1984-present, as officially reported to ICES. (Working Group figures are given in Table 9.2).

| Country | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 | 4 | 5 | 10 | 12 | 4 | 4 | 1 | 8 | 18 |
| France | 38 | 31 | 39 | 50 | 47 | n/a | n/a | n/a | 73 | 41 |
| Ireland | 199 | 341 | 275 | 797 | 363 | 215 | 80 | 254 | 251 | 252 |
| Netherlands | - | - | - | - | - | - | - | - | - | - |
| UK(E\&W) ${ }^{1}$ | 29 | 28 | 22 | 41 | 74 | 252 | 177 | 204 | 244 | 260 |
| UK (Isle of Man) | 2 | 5 | 4 | 3 | 3 | 3 | 5 | 14 | 13 | 19 |
| UK (N. Ireland) | 38 | 215 | 358 | 230 | 196 | ... | ... | ... | ... | ... |
| UK (Scotland) | 78 | 104 | 23 | 156 | 52 | 86 | 316 | 143 | 114 | 140 |
| Total | 387 | 728 | 726 | 1,287 | 747 | 560 | 582 | 616 | 703 | 730 |
| Country | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| Belgium | 22 | 32 | 34 | 55 | 104 | 53 | 22 | 68 | 44 | 20 |
| France | 22 | 58 | 105 | 74 | 86 | n/a | 49 | 184 | 72 | 146 |
| Ireland | 246 | 320 | 798 | 1,005 | 1,699 | 759 | 1,238 | 652 | 401 | 229 |
| Netherlands | - | - | 1 | 14 | 10 | 5 | 2 | - | - | - |
| UK(E\&W) ${ }^{1}$ | 301 | 294 | 463 | 717 | 1,023 | 1,479 | 1,061 | 1,238 | 551 | 248 |
| UK (Isle of Man) | 24 | 27 | 38 | 9 | 13 | 7 | 19 | 1 | - | - |
| UK (N. Ireland) | ... | ... | ... | ... | ... | ... | ... | ... | ... | ... |
| UK (Scotland) | 66 | 110 | 14 | 51 | 80 | 67 | 56 | 86 | 47 | 31 |
| Total | 681 | 841 | 1,453 | 1,925 | 3,015 | 2,370 | 2,447 | 2,229 | 1,115 | 674 |


| Country | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 15 | 22 | 23 | 30 | 15 | 7 | 9 | 16 | 13 | 6.2 |
| France | 20 | 36 | 20 | 11 | 6 | 3 | 2 | 8 | 3 | . 7 |
| Ireland | 296 | 139 | 184 | 477 | 319 | 388 | 333 | 434 | 561 | 492 |
| Netherlands | - | - |  | - | - | - | - | - | - | - |
| UK (England \& Wales) ${ }^{1}$ | 421 | 344 | 419 | 559 | 521 | 446 | 593 | 355 | 236 | 154 |
| UK (Isle of Man) | - | - | - | - | 1 | 1 | - | - | <1 | <. 1 |
| UK (N. Ireland) | $\ldots$ | ... | ... | ... | ... | ... | ... | ... | ... | ... |
| UK (Scotland) | 9 | 6 | 9 | 1 | 17 | 1 | 2 |  |  | - |
| United Kingdom |  |  |  |  |  |  |  |  | 236 | 154 |
| Total | 761 | 547 | 655 | 1078 | 879 | 846 | 939 | 813 | 813 | 654 |


| Country | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020* | 2021* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 7 | 7 | 5 | 5 | 4 | 9 | 4 | 3 |
| France | 0 | 7 | 1 | 5 | 0 | 0 | 0 | 0 |
| Ireland | 541 | 507 | 632 | 114 | 949 | 1347 | 754 | 1162 |
| Netherlands | - | - |  | - | - | - | - | - |
| UK (England \& Wales) ${ }^{1}$ | - | - |  | - | - | - | - | - |
| UK (Isle of Man) | <1 | <1 |  | - | - | - | - | - |
| UK (N. Ireland) | $\cdots$ | - |  | - | - | - | - | - |
| UK (Scotland) | - | - |  | - | - | - | - | - |
| United Kingdom | 426 | 634 | 825 | 1240 | 1580 | 1197 | 539 | 884 |
| Total | 974 | 1154 | 1463 | 2363 | 2532 | 2553 | 1296 | 2048 |

* Preliminary.
${ }^{1}$ 1989-2015 Northern Ireland included with England and Wales.
n/a = not available.

Table 9.2. Haddock in 7.a. Total international landings of haddock from the Irish Sea, 1972-present as officially reported to ICES. Working Group figures, assuming 1972-1992 official landings to be correct, are also given. The 1993-2005 WG estimates include sampled-based estimates of landings at a number of Irish Sea ports. Sample-based evidence confirms more accurate catch reporting since 2006. Landings in tonnes live weight. Since 1993 the landings have been corrected to exclude catches from the southernmost rectangles, which are not considered part of this stock.
$\left.\begin{array}{lccccc}\hline \text { Year } & \begin{array}{l}\text { Official land- } \\ \text { ings }\end{array} & \begin{array}{l}\text { WG land- } \\ \text { ings }\end{array} & \begin{array}{l}\text { ICES dis- } \\ \text { cards* }\end{array} & \begin{array}{l}\text { ICES } \\ \text { catch }\end{array} & \text { \% Discard }\end{array} \begin{array}{l}\text { Landings taken or reported in rectan- } \\ \text { gles 33E2 and 33E3 }\end{array}\right]$

| Year | Official landings | WG landings | ICES discards** | ICES <br> catch | \% Discard | Landings taken or reported in rectangles 33E2 and 33E3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 3015 | 4874 | 1015 | 5889 | 17\% | 28 |
| 1999 | 2370 | 4095 | 634 | 4729 | 13\% | 34 |
| 2000 | 2447 | 1357 | 802 | 2159 | 37\% | 11 |
| 2001 | 2229 | 2246 | 269 | 2515 | 11\% | 74 |
| 2002 | 1115 | 1817 | 387 | 2204 | 18\% | 82 |
| 2003 | 674 | 659 | - | - | - | 64 |
| 2004 | 761 | 1217 | 392 | 1609 | 24\% | 53 |
| 2005 | 547 | 666 | 551 | 1217 | 45\% | 35 |
| 2006 | 655 | 633 | 306 | 939 | 33\% | 26 |
| 2007 | 1078 | 886 | 722 | 1608 | 45\% | 222 |
| 2008 | 879 | 786 | 643 | 1429 | 45\% | 194 |
| 2009 | 846 | 581 | 579 | 1160 | 50\% | 285 |
| 2010 | 939 | 679 | 508 | 1187 | 43\% | 267 |
| 2011 | 813 | 446 | 307 | 753 | 41\% | 374 |
| 2012 | n/a | 343 | 599 | 942 | 64\% | 473 |
| 2013 | 654 | 254 | 283 | 537 | 53\% | 410 |
| 2014 | 953 | 518 | 488 | 1006 | 49\% | 444 |
| 2015 | 1154 | 833 | 652 | 1451 | 44\% | 322 |
| 2016 | 1463 | 1008 | 298 | 1306 | 23\% | 455 |
| 2017 | 2363 | 1662 | 333 | 1995 | 17\% | 715 |
| 2018 | 2532 | 1993 | 568 | 2561 | 22\% | 532 |
| 2019 | 2537 | 1778 | 672 | 2450 | 27\% | 759 |
| 2020 | 1296 | 7423 | 234 | 976 | 24\% | 554 |
| 2021 | 2048 | 1219 | 674 | 1891 | 36\% | 827 |

Table 9.3. Haddock in 7.a: stock weights-at-age.

|  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 |
| 1993 | 0.02 | 0.095 | 0.42 | 1.043 | 1.759 | 2.563 |
| 1994 | 0.02 | 0.083 | 0.338 | 0.968 | 1.999 | 3.028 |
| 1995 | 0.02 | 0.085 | 0.347 | 0.785 | 1.708 | 3.219 |
| 1996 | 0.02 | 0.083 | 0.359 | 0.788 | 1.319 | 2.718 |
| 1997 | 0.022 | 0.07 | 0.357 | 0.863 | 1.435 | 2.391 |
| 1998 | 0.018 | 0.06 | 0.253 | 0.743 | 1.384 | 2.165 |
| 1999 | 0.016 | 0.057 | 0.226 | 0.561 | 1.294 | 2.262 |
| 2000 | 0.017 | 0.048 | 0.23 | 0.51 | 0.966 | 2.123 |
| 2001 | 0.018 | 0.051 | 0.201 | 0.548 | 0.93 | 1.822 |
| 2002 | 0.017 | 0.056 | 0.215 | 0.472 | 0.983 | 1.637 |
| 2003 | 0.017 | 0.05 | 0.229 | 0.485 | 0.798 | 1.52 |
| 2004 | 0.017 | 0.041 | 0.199 | 0.509 | 0.816 | 1.306 |
| 2005 | 0.018 | 0.031 | 0.165 | 0.459 | 0.902 | 1.347 |
| 2006 | 0.014 | 0.033 | 0.128 | 0.378 | 0.803 | 1.435 |
| 2007 | 0.019 | 0.034 | 0.136 | 0.299 | 0.68 | 1.402 |
| 2008 | 0.014 | 0.037 | 0.139 | 0.31 | 0.515 | 1.167 |
| 2009 | 0.025 | 0.042 | 0.153 | 0.326 | 0.563 | 0.98 |
| 2010 | 0.017 | 0.04 | 0.176 | 0.357 | 0.58 | 0.945 |
| 2011 | 0.018 | 0.052 | 0.167 | 0.407 | 0.624 | 0.937 |
| 2012 | 0.012 | 0.057 | 0.209 | 0.375 | 0.688 | 0.96 |
| 2013 | 0.023 | 0.059 | 0.233 | 0.491 | 0.673 | 1.115 |
| 2014 | 0.022 | 0.038 | 0.238 | 0.512 | 0.812 | 1.04 |
| 2015 | 0.017 | 0.046 | 0.153 | 0.577 | 0.97 | 1.371 |
| 2016 | 0.021 | 0.047 | 0.192 | 0.354 | 1.015 | 1.533 |
| 2017 | 0.022 | 0.054 | 0.137 | 0.347 | 0.809 | 1.476 |
| 2018 | 0.023 | 0.068 | 0.196 | 0.472 | 0.601 | 0.987 |
| 2019 | 0.024 | 0.066 | 0.121 | 0.480 | 0.636 | 1.04 |
| 2020* | 0.023 | 0.063 | 0.151 | 0.433 | 0.682 | 1.168 |
| 2021 | 0.034 | 0.064 | 0.117 | 0.372 | 0.552 | 0.967 |

[^4]Table 9.4. Haddock in 7.a: Catch numbers-at-age.

|  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 |
| 1993 | 790 | 1568 | 2066 | 19 | 1 | 1 |
| 1994 | 16857 | 821 | 258 | 922 | 3 | 2 |
| 1995 | 950 | 8079 | 1587 | 107 | 220 | 5 |
| 1996 | 15171 | 1380 | 5510 | 728 | 16 | 30 |
| 1997 | 347 | 8828 | 1528 | 2388 | 201 | 16 |
| 1998 | 4209 | 4642 | 10532 | 252 | 488 | 42 |
| 1999 | 4944 | 3200 | 3436 | 4773 | 25 | 57 |
| 2000 | 287 | 11118 | 1771 | 466 | 457 | 418 |
| 2001 | 7883 | 425 | 3246 | 1074 | 30 | 89 |
| 2002 | 2105 | 8229 | 789 | 2063 | 142 | 18 |
| 2003 | 2000 | 2000 | 400 | 800 | 50 | 25 |
| 2004 | 10797 | 2056 | 421 | 827 | 46 | 78 |
| 2005 | 6048 | 4342 | 1416 | 285 | 193 | 34 |
| 2006 | 5334 | 2971 | 656 | 524 | 63 | 51 |
| 2007 | 2282 | 3537 | 3371 | 671 | 60 | 47 |
| 2008 | 2158 | 4569 | 2052 | 837 | 242 | 36 |
| 2009 | 4327 | 2490 | 2021 | 629 | 121 | 36 |
| 2010 | 3933 | 4058 | 834 | 464 | 309 | 59 |
| 2011 | 5669 | 2324 | 942 | 239 | 97 | 52 |
| 2012 | 6235 | 2799 | 774 | 201 | 27 | 28 |
| 2013 | 4525 | 1162 | 558 | 156 | 41 | 17 |
| 2014 | 1392 | 3854 | 1265 | 189 | 17 | 10 |
| 2015 | 518 | 1915 | 3087 | 324 | 63 | 5 |
| 2016 | 512 | 1845 | 907 | 1079 | 109 | 108 |
| 2017 | 231 | 783 | 2234 | 829 | 1096 | 78 |
| 2018 | 56 | 1039 | 5325 | 2845 | 426 | 526 |
| 2019 | 415 | 5276 | 4528 | 604 | 1132 | 467 |
| 2020 | 0 | 3269 | 559 | 282 | 598 | 367 |
| 2021 | 716 | 782 | 3064 | 2256 | 364 | 133 |

Table 9.5. Haddock in 7.a: Available tuning data and maturity ogive
IRISH SEA haddock, 2013 WG,ANON, COMBSEX, TUNING DATA(effort, nos-at-age)
101
NIGFS-WIBTS-Q1
19932021
110.210 .25

05

| 1 | 0 | 139 | 569 | 31 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 644 | 58 | 183 | 0 | 0 |
| 1 | 0 | 24823 | 437 | 0.1 | 43 | 0 |
| 1 | 0 | 1065 | 3743 | 67 | 3 | 1.1 |
| 1 | 0 | 25118 | 474 | 1457 | 44 | 2.1 |
| 1 | 0 | 3913 | 8694 | 70 | 105 | 1.1 |
| 1 | 0 | 6058 | 680 | 2072 | 16 | 11.1 |
| 1 | 0 | 14028 | 1853 | 64 | 147 | 5 |
| 1 | 0 | 3277 | 6990 | 770 | 40 | 20.1 |
| 1 | 0 | 28755 | 842 | 1059 | 78 | 1.1 |
| 1 | 0 | 6966 | 14162 | 341 | 356 | 26.1 |
| 1 | 0 | 19945 | 2379 | 2206 | 45 | 35.1 |
| 1 | 0 | 24488 | 6454 | 406 | 234 | 15 |
| 1 | 0 | 13444 | 12721 | 2194 | 91 | 33.1 |
| 1 | 0 | 20918 | 11325 | 3661 | 240 | 27 |
| 1 | 0 | 7480 | 12009 | 2559 | 495 | 48.1 |
| 1 | 0 | 9345 | 3888 | 2877 | 163 | 42 |
| 1 | 0 | 17058 | 1765 | 524 | 239 | 27 |
| 1 | 0 | 17278 | 5543 | 299 | 67 | 50 |
| 1 | 0 | 13509 | 5266 | 1095 | 38 | 13 |
| 1 | 0 | 8245 | 5202 | 751 | 119 | 20 |
| 1 | 0 | 33807 | 2260 | 773 | 108 | 22 |
| 1 | 0 | 15495 | 22420 | 1297 | 407 | 44 |
| 1 | 0 | 14418 | 9109 | 5594 | 205 | 38 |
| 1 | 0 | 4321 | 18887 | 5524 | 323 | 33 |
| 1 | 0 | 7897 | 4683 | 7086 | 1709 | 1369 |
| 1 | 0 | 38570 | 6789 | 814 | 832 | 183 |
| 1 | 0 | 16709 | 28889 | 2571 | 260 | 257 |
| 1 | 0 | 2478.3 | 17390.6 | 6690.7 | 550.6 | 41 |


| $\begin{aligned} & \text { NIG } \\ & 199 \\ & 1 \\ & 1 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { BTS-Q4 } \\ & 1 \begin{array}{r} 1 \\ \hline \end{array} \mathbf{8} 8 \\ & \hline \end{aligned}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 36.127 | 0.716 | 3.965 | 0 | 0 |
| 1 | 2.042 | 151.766 | 1.171 | 0.959 | 0 |
| 1 | 15.289 | 101.536 | 0.753 | 0 | 0.045 |
| 1 | 1067.99 | 13.327 | 13.2 | 0.092 | 0.001 |
| 1 | 160.434 | 398.722 | 1.81 | 0.886 | 0.04 |
| 1 | 365.679 | 10.521 | 39.889 | 0.08 | 0.034 |
| 1 | 685.913 | 28.002 | 0.527 | 1.633 | 0.001 |
| 1 | 59.867 | 93.66 | 5.533 | 0.125 | 0.104 |
| 1 | 584.902 | 19.354 | 28.408 | 0.947 | 0 |
| 1 | 146.491 | 105.115 | 1.18 | 3.372 | 0 |
| 1 | 552.309 | 59.354 | 30.746 | 0.295 | 0.27 |
| 1 | 666.652 | 167.224 | 7.422 | 4.911 | 0.001 |
| 1 | 476.2 | 122.094 | 12.378 | 0.264 | 0.052 |
| 1 | 387.556 | 111.692 | 35.717 | 2.228 | 0.441 |
| 1 | 94.667 | 102.086 | 37.1 | 11.654 | 0.375 |
| 1 | 88.61 | 46.338 | 23.832 | 1.991 | 0.33 |
| 1 | 451.303 | 45.695 | 6.139 | 4.891 | 0.23 |
| 1 | 219.533 | 82.392 | 5.858 | 1.752 | 0.973 |
| 1 | 207.925 | 42.145 | 7.808 | 1.044 | 0.093 |
| 1 | 165.294 | 79.593 | 12.05 | 1.275 | 0 |
| 1 | 1004.22 | 8.279 | 1.531 | 0.179 | 0 |
| 1 | 339.218 | 311.607 | 68.768 | 3.016 | 0.423 |
| 1 | 455.385 | 81.189 | 108.663 | 2.309 | 0.362 |
| 1 | 99.046 | 154.865 | 52.207 | 4.273 | 0.281 |
| 1 | 191.946 | 42.885 | 90.324 | 15.934 | 6.202 |
| 1 | 690.663 | 167.338 | 12.891 | 16.507 | 2.003 |
| 1 | 21.174 | 179.518 | 169.383 | 8.19 | 0.58 |
| 1 | 133.3 | 1209 | 50 | 13.2 | 0.66 |


| NIMIK |  |
| :---: | :---: |
| 19942021 |  |
| 110.380 .47 |  |
| 00 |  |
| 1 | 47000 |
| 1 | 1700 |
| 1 | 47800 |
| 1 | 14500 |
| 1 | 2500 |
| 1 | 15400 |
| 1 | 1700 |
| 1 | 17100 |
| 1 | 1200 |
| 1 | 4250 |
| 1 | 25970 |
| 1 | 8250 |
| 1 | 40240 |
| 1 | 3820 |
| 1 | 6638 |
| 1 | 18540 |
| 1 | 4532 |
| 1 | 6606 |
| 1 | 9818 |
| 1 | 28325 |
| 1 | 12892 |
| 1 | 48463 |
| 1 | 1800 |
| 1 | 26900 |
| 1 | 30954 |
| 1 | 23942 |
| 1 | NA |
| 1 | 16800 |



Maturity ogive at-age.

| Year | 0 | 1 | 2 | 3 | 4 | 5+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 0 | 0.762 | 0.99 | 1 | 1 |
| 1994 | 0 | 0 | 0.762 | 0.99 | 1 | 1 |
| 1995 | 0 | 0 | 0.784 | 0.99 | 1 | 1 |
| 1996 | 0 | 0 | 0.78 | 0.99 | 1 | 1 |
| 1997 | 0 | 0 | 0.777 | 0.99 | 1 | 1 |
| 1998 | 0 | 0 | 0.775 | 0.99 | 1 | 1 |
| 1999 | 0 | 0 | 0.773 | 0.99 | 1 | 1 |
| 2000 | 0 | 0 | 0.771 | 0.99 | 1 | 1 |
| 2001 | 0 | 0 | 0.769 | 0.99 | 1 | 1 |
| 2002 | 0 | 0 | 0.767 | 0.99 | 1 | 1 |
| 2003 | 0 | 0 | 0.763 | 0.99 | 1 | 1 |
| 2004 | 0 | 0 | 0.762 | 0.99 | 1 | 1 |
| 2005 | 0 | 0 | 0.771 | 0.99 | 1 | 1 |
| 2006 | 0 | 0 | 0.784 | 0.99 | 1 | 1 |
| 2007 | 0 | 0 | 0.797 | 0.99 | 1 | 1 |
| 2008 | 0 | 0.01 | 0.809 | 0.99 | 1 | 1 |
| 2009 | 0 | 0.01 | 0.817 | 0.99 | 1 | 1 |
| 2010 | 0 | 0.01 | 0.825 | 0.99 | 1 | 1 |
| 2011 | 0 | 0.01 | 0.833 | 0.99 | 1 | 1 |
| 2012 | 0 | 0.01 | 0.841 | 0.99 | 1 | 1 |
| 2013 | 0 | 0.02 | 0.847 | 0.99 | 1 | 1 |
| 2014 | 0 | 0.02 | 0.846 | 0.99 | 1 | 1 |
| 2015 | 0 | 0.02 | 0.848 | 1 | 1 | 1 |
| 2016 | 0 | 0.03 | 0.85 | 1 | 1 | 1 |
| 2017 | 0 | 0.03 | 0.851 | 1 | 1 | 1 |
| 2018 | 0 | 0.03 | 0.853 | 1 | 1 | 1 |
| 2019 | 0 | 0.03 | 0.853 | 1 | 1 | 1 |
| 2020 | 0 | 0.03 | 0.854 | 1 | 1 | 1 |
| 2021 | 0 | 0.06 | 0.855 | 1 | 1 | 1 |

Table 9.6. Haddock in 7.a: F-at-age.

|  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 |
| 1993 | 0.032533 | 0.3892 | 0.861503 | 0.91716 | 0.920132 | 0.920281 |
| 1994 | 0.04169 | 0.498745 | 1.103981 | 1.175304 | 1.179113 | 1.179304 |
| 1995 | 0.038902 | 0.465392 | 1.030154 | 1.096706 | 1.100261 | 1.100439 |
| 1996 | 0.02542 | 0.304103 | 0.673138 | 0.716625 | 0.718948 | 0.719065 |
| 1997 | 0.028486 | 0.340777 | 0.754316 | 0.803049 | 0.805651 | 0.805782 |
| 1998 | 0.032306 | 0.386486 | 0.855495 | 0.910764 | 0.913716 | 0.913864 |
| 1999 | 0.045476 | 0.544036 | 1.204235 | 1.282034 | 1.28619 | 1.286398 |
| 2000 | 0.03025 | 0.361882 | 0.801032 | 0.852782 | 0.855546 | 0.855685 |
| 2001 | 0.123163 | 0.404643 | 0.73249 | 0.786004 | 0.551934 | 0.393002 |
| 2002 | 0.154348 | 0.507099 | 0.917957 | 0.985021 | 0.691685 | 0.492511 |
| 2003 | 0.126182 | 0.414562 | 0.750445 | 0.805271 | 0.565464 | 0.402636 |
| 2004 | 0.122492 | 0.402438 | 0.728499 | 0.781721 | 0.548927 | 0.390861 |
| 2005 | 0.100897 | 0.331491 | 0.600069 | 0.643908 | 0.452154 | 0.321954 |
| 2006 | 0.057887 | 0.190183 | 0.344271 | 0.369423 | 0.25941 | 0.184711 |
| 2007 | 0.096329 | 0.316482 | 0.572899 | 0.614753 | 0.431682 | 0.307377 |
| 2008 | 0.144765 | 0.505022 | 0.575852 | 0.547062 | 0.299876 | 0.145633 |
| 2009 | 0.112144 | 0.391222 | 0.446092 | 0.423789 | 0.232303 | 0.112817 |
| 2010 | 0.163635 | 0.570852 | 0.650915 | 0.618372 | 0.338965 | 0.164617 |
| 2011 | 0.088326 | 0.308132 | 0.351348 | 0.333782 | 0.182965 | 0.088856 |
| 2012 | 0.093186 | 0.325087 | 0.370681 | 0.352148 | 0.193032 | 0.093745 |
| 2013 | 0.007748 | 0.06819 | 0.120837 | 0.120837 | 0.120837 | 0.120837 |
| 2014 | 0.01022 | 0.089946 | 0.159389 | 0.159388 | 0.159389 | 0.159389 |
| 2015 | 0.009024 | 0.079421 | 0.140738 | 0.140738 | 0.140738 | 0.140738 |
| 2016 | 0.006105 | 0.053732 | 0.095216 | 0.095216 | 0.095216 | 0.095216 |
| 2017 | 0.007814 | 0.068769 | 0.121862 | 0.121862 | 0.121862 | 0.121862 |
| 2018 | 0.009921 | 0.087311 | 0.15472 | 0.15472 | 0.15472 | 0.15472 |
| 2019 | 0.010692 | 0.094097 | 0.166745 | 0.166745 | 0.166745 | 0.166745 |
| 2020 | 0.004897 | 0.043096 | 0.076369 | 0.076369 | 0.076369 | 0.076369 |
| 2021 | 0.010063 | 0.088559 | 0.156931 | 0.156931 | 0.156931 | 0.156931 |

Table 9.7. Forecast input data.

| Variable | Value | Source | Notes |
| :---: | :---: | :---: | :---: |
| F ages 2-4 (2021) | 0.159 | ICES (2022a) | Fsq $=\mathrm{F}_{\text {average(2018-2021) }}$ excluding 2020 |
| SSB (2023) | 11817 | ICES (2022a) | Short-term forecast |
| $R$ age 0 (2022 and 2023) (thousand) | 364084 | ICES (2022a) | Geometric mean (1993-2019) |
| Catch (2022) | 1846 | ICES (2022a) | Short-term forecast, fishing at $\mathrm{F}_{\text {sq }}$ |
| Wanted catch * (2022) | 1545 | ICES (2022a) | Average discard rate (2019-2021) |
| Unwanted catch *(2022) | 301 | ICES (2022a) | Average discard rate (2019-2021) |

* "Wanted catch" is used to describe fish that would be landed in the absence of the EU landing obligation.

Table 9.8. Haddock in Division 7.a. Assessment summary. All weights are in tonnes, recruitment (age 0) in thousands. Low and high refer to $95 \%$ confidence intervals.

| Year | Recruitment age 0 |  |  | SSB |  |  | Landings | Discards* | F ages 2-4 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Value | High | Low | Value | High |  |  | Low | Value | High |
| 1993 | 116600 | 152730 | 188861 | 1640 | 2288 | 2937 | 813 | 365 | 0.40 | 0.68 | 0.96 |
| 1994 | 421898 | 520923 | 619948 | 1408 | 2161 | 2913 | 1042 | 468 | 0.40 | 0.72 | 1.03 |
| 1995 | 39534 | 63565 | 87597 | 1492 | 2312 | 3131 | 1736 | 780 | 0.52 | 0.94 | 1.37 |
| 1996 | 1087922 | 1340978 | 1594034 | 3603 | 4773 | 5944 | 2981 | 709 | 0.48 | 0.75 | 1.02 |
| 1997 | 149800 | 210066 | 270332 | 2638 | 3952 | 5265 | 3547 | 895 | 0.57 | 0.94 | 1.31 |
| 1998 | 260460 | 342308 | 424155 | 6375 | 8051 | 9727 | 4874 | 1015 | 0.68 | 0.98 | 1.27 |
| 1999 | 539265 | 669483 | 799702 | 4062 | 5504 | 6946 | 4095 | 634 | 0.97 | 1.46 | 1.95 |
| 2000 | 64449 | 98739 | 133030 | 1769 | 2646 | 3522 | 1357 | 802 | 0.60 | 1.04 | 1.48 |
| 2001 | 553816 | 698136 | 842457 | 2689 | 3746 | 4803 | 2246 | 269 | 0.48 | 0.73 | 0.98 |
| 2002 | 91899 | 132937 | 173975 | 1840 | 2790 | 3741 | 1817 | 387 | 0.59 | 0.95 | 1.31 |
| 2003 | 309707 | 419741 | 529776 | 2224 | 3233 | 4243 | 1517 | 390 | 0.46 | 0.77 | 1.09 |
| 2004 | 500341 | 642619 | 784897 | 1418 | 2371 | 3324 | 1217 | 392 | 0.43 | 0.76 | 1.09 |
| 2005 | 384571 | 490288 | 596004 | 1389 | 2229 | 3070 | 666 | 551 | 0.35 | 0.63 | 0.90 |
| 2006 | 450802 | 558138 | 665473 | 1968 | 2885 | 3803 | 633 | 306 | 0.198 | 0.35 | 0.51 |
| 2007 | 169235 | 219870 | 270505 | 2828 | 3859 | 4889 | 886 | 722 | 0.39 | 0.59 | 0.80 |
| 2008 | 115259 | 154177 | 193095 | 2834 | 3931 | 5027 | 786 | 643 | 0.34 | 0.52 | 0.71 |
| 2009 | 256751 | 328391 | 400031 | 2211 | 3343 | 4475 | 581 | 579 | 0.25 | 0.40 | 0.55 |
| 2010 | 186107 | 242003 | 297898 | 1804 | 2862 | 3920 | 679 | 508 | 0.37 | 0.61 | 0.84 |
| 2011 | 229755 | 297411 | 365066 | 1538 | 2539 | 3540 | 446 | 307 | 0.189 | 0.32 | 0.45 |
| 2012 | 210728 | 286718 | 362707 | 1776 | 2823 | 3869 | 343 | 599 | 0.197 | 0.34 | 0.47 |
| 2013 | 1092158 | 1375162 | 1658166 | 2315 | 3638 | 4960 | 254 | 282 | 0.076 | 0.133 | 0.191 |
| 2014 | 475004 | 626559 | 778114 | 3470 | 5107 | 6744 | 518 | 488 | 0.103 | 0.172 | 0.24 |
| 2015 | 706489 | 927283 | 1148077 | 8046 | 10842 | 13637 | 833 | 652 | 0.092 | 0.150 | 0.21 |
| 2016 | 207449 | 293473 | 379497 | 10782 | 14457 | 18131 | 1008 | 298 | 0.062 | 0.101 | 0.139 |
| 2017 | 273576 | 383175 | 492774 | 13817 | 18505 | 23192 | 1662 | 333 | 0.079 | 0.129 | 0.178 |
| 2018 | 632164 | 872205 | 1112247 | 13986 | 18988 | 23990 | 1993 | 568 | 0.097 | 0.163 | 0.23 |
| 2019 | 389547 | 570827 | 752108 | 11121 | 15753 | 20385 | 1778 | 672 | 0.101 | 0.175 | 0.25 |
| 2020 | 25337 | 61526 | 97716 | 10081 | 14579 | 19078 | 742 | 177 | 0.042 | 0.079 | 0.116 |
| 2021 | 176790 | 321561 | 466333 | 10297 | 14944 | 19590 | 1219 | 672 | 0.088 | 0.161 | 0.23 |
| 2022 |  | 370456** |  |  | 14274 |  |  |  |  |  |  |

[^5]Table 9.9. Haddock in Division 7.a. Annual catch scenarios. All weights are in tonnes.

| Basis | Total catch (2023) | Wanted catch* (2023) | Unwanted catch* (2023) | $\begin{aligned} & \mathrm{F}_{\text {total }} \\ & (2023) \end{aligned}$ | $\begin{aligned} & F_{\text {wanted }} \\ & (2023) \end{aligned}$ | $F_{\text {un- }}$ wanted (2023) | $\begin{aligned} & \text { SSB } \\ & (2024) \end{aligned}$ | \%SSB change ** | \%Ad- <br> vice <br> change <br> $\wedge$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice Basis |  |  |  |  |  |  |  |  |  |
| EU MAP ${ }^{* * *}$ : $\mathrm{F}_{\text {MSY }}$ | 2648 | 2107 | 541 | 0.28 | 0.171 | 0.109 | 9321 | -21 | -12.8 |
| F = MAP $\mathrm{F}_{\text {MSY lower }}$ | 1956 | 1560 | 396 | 0.2 | 0.122 | 0.078 | 10044 | -15 | -35.6 |
| F = MAP $\mathrm{F}_{\text {MSY upper }}$ | 3216 | 2554 | 661 | 0.35 | 0.21 | 0.137 | 8732 | -26 | 5.9 |
| Other scenarios |  |  |  |  |  |  |  |  |  |
| $\mathrm{F}=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 12115 | 2.5 | -100 |
| $\mathrm{F}=\mathrm{F}_{\mathrm{pa}}$ | 3676 | 2915 | 761 | 0.41 | 0.25 | 0.16 | 8258 | -30 | 21 |
| $\mathrm{F}=\mathrm{F}_{\text {lim }}$ | 4323 | 3421 | 903 | 0.5 | 0.3 | 0.195 | 7597 | -36 | 42.3 |
| F = F2022 | 1587 | 1267 | 320 | 0.15947 | 0.097 | 0.062 | 10432 | -11.7 | -47.8 |
| SSB2023 $=\mathrm{B}_{\text {lim }}$ | 9125 | 7029 | 2095 | 1.52547 | 0.93 | 0.6 | 2994 | -75 | 200 |
| SSB2023 $=\mathrm{B}_{\text {pa }}$ | 7834 | 6092 | 1742 | 1.15805 | 0.71 | 0.45 | 4160 | -65 | 158 |
| SSB2023=MSY |  |  |  |  |  |  |  |  |  |
| $\mathrm{B}_{\text {trigger }}$ | 7704 | 5997 | 1708 | 1.12632 | 0.69 | 0.44 | 4281 | -64 | 154 |
| SSB2023=SSB2022 | 280 | 224 | 56 | 0.02653 | 0.0162 | 0.0104 | 11817 | 0 | -90.8 |

* "Wanted" and "unwanted" catch are used to describe fish that would be landed and discarded in the absence of the EU landing obligation, based on discard rate estimates for 2019-2021.
** SSB 2024 relative to SSB 2023.
*** EU multiannual plan (MAP) for the Western Waters (EU, 2019).
${ }^{\wedge}$ Advice value for 2023 relative to the Fmš advice value for 2022 ( 3038 tonnes).

Table 9.10. Haddock in 7.a Management reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $B_{\text {trigger }}$ | 4281 tonnes | 5th percentile of BMSY; Irish Sea haddock has been fished at, or below $F_{\text {MSY }}$ for >five years. | ICES <br> (2018a) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.28 | Median point estimates of EqSim with segmented regression stock-recruitment relationship | ICES <br> (2018a) |
|  | $\mathrm{F}_{\text {MSYLower }}$ | 0.20 | F at 95\% of MSY below $\mathrm{F}_{\text {MSY }}$ | ICES <br> (2018a) |
|  | $\mathrm{F}_{\text {MSYUpper }}$ | 0.35 | F at $95 \%$ of MSY above $\mathrm{F}_{\text {MSY }}$ | ICES <br> (2018a) |
| Precautionary approach | $\mathrm{B}_{\text {lim }}$ | 2994 tonnes | Lowest observed SSB with >75th percentile recruitment | ICES <br> (2018a) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 4160 tonnes | $\mathrm{B}_{\text {lim }}$ combined with the assessment error; $\mathrm{B}_{\mathrm{lim}} \times$ $\exp (1.645 \times \sigma) ; \sigma=0.20$ | ICES <br> (2018a) |
|  | $\mathrm{F}_{\text {lim }}$ | 0.50 | F with $50 \%$ probability of SSB $<\mathrm{B}_{\text {lim }}$ | ICES <br> (2018a) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.41 | $\mathrm{F}_{\mathrm{p} 0.05}$; the F that leads to $\mathrm{SSB}>\mathrm{B}_{\text {lim }}$ with $95 \%$ probability | ICES <br> (2018a) |
| Management plan | SSB ${ }_{\text {MGT }}$ | Not applicable |  |  |
|  | $\mathrm{F}_{\text {MGT }}$ | Not applicable |  |  |



Figure 9.1. Haddock in 7.a: Growth of haddock in the Irish Sea. Top two panels: mean length-at-age in UK(NI) groundfish surveys in March (NIGFS-WIBTS-Q1), by year and age, and expected mean weight-at-length based on length-weight parameters from each survey. Lower panels: mean length-at-age from March surveys, and from Quarter 1 commercial landings at-age 3 and over, by year class. Lines are von Bertalanffy model fits with year-class effect included. Model residuals are shown for the fit without year-class effects, and for the fit with year-class effects.


Figure 9.2. Haddock in 7.a: Trends in log-standardised survey indices.

## UKFspW


log index

NIGFS-WIBTS-Q1


NIGFS-WIBTS-Q4

log index

Figure 9.3. Haddock in 7.a: Scatterplot matrix of log indices of cohorts at different ages.


Figure 9.4. Standardised residuals from fitted and observed catch age proportions.


Figure 9.5. Fitted and observed catch from update assessment.


Figure 9.6. Fitted and observed catch age proportions from update assessment.


Figure 9.7. Observed catch numbers 2008-present.


Figure 9.8. Fitted and observed index series from update assessment.


Figure 9.9. Fitted and observed index age proportions from update assessment.



$F_{\text {bar }} 0.082$

Figure 9.10. Retrospective plot the final update model.


Figure 9.11. Haddock in Division7.a. Historical assessment results.


Figure 9.12. Haddock in 7a. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.

## 10 Megrim (Lepidorhombus ssp.) in divisions 4.a and 6.a (northern North Sea, West of Scotland)

## Type of assessment in 2022

Update of 2021 assessment with new landings and survey data. The model used to carry out the assessment is the Schaefer Surplus production process model in R and Winbugs.

## ICES advice applicable to 2023

ICES advise that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, catches in 2023 that correspond to the F ranges in the plan are between 5550 tonnes and 7200 tonnes.

### 10.1 General

## Stock description and management units

Megrim stock structure is uncertain and historically the Working Group has considered megrim populations in $6 . a$ and $6 . \mathrm{b}$ as separate stocks. The review group questioned the basis for this in 2004. Data collected during an EC study contract (98/096) on the 'Distribution and biology of anglerfish and megrim in the waters to the West of Scotland' showed significantly different growth parameters and significant population structure difference between megrim sampled in 6.a and 6.b (Anon, 2001). Spawning fish occur in both areas but whether these populations are reproductively isolated is not clear. As noted by WGNSDS (2008), megrim in 4.a has historically not been considered by ICES. Since 2009 data from 4 and 2.a are included in this report, but international catch and weight-at-age data for 4 , prior 2006 were not available to the working group. Given that there is little evidence to suggest that megrim in $6 . a$ and $4 . a$ are separate stocks, based on a visual inspection of the spatial distribution of commercial landings and fishery-independent survey data, WKFLAT (2011) concluded that megrim in 6 .a and 4 .a should be considered as a single stock. This has subsequently been supported through recent genetic studies (MacDonald and Prieto, 2012) indicating that there is one stock consisting of divisions 4.a (northern North Sea) and 6.a (West of Scotland) and another separate stock in Division 6.b (Rockall).


TAC/Management area
$\square$ Assessment area

## Management area (red boxes) and assessment area (blue hatched boxes).

| Species: | Megrims <br> Lepidorhombus spp. |  |  | Zone: | United Kingdom and Union waters of 4; United Kingdom waters of 2a <br> (LEZ/2AC4-C) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 8 | ${ }^{(2)}$ | Analy |  |
| Denmark |  | 7 | (2) | Artic | his Regulation applies |
| Germany |  | 7 | (2) |  |  |
| France |  | 45 | (a) |  |  |
| Netherlands |  | 36 | (2) |  |  |
| Union |  | 103 | (2) |  |  |
| United Kingdom |  | 2660 | (2) |  |  |
| TAC |  | 2763 |  |  |  |

(a) Special condition: of which up to $20 \%$ may be fished in United Kingdom, Union and international waters of 6a north of $58^{\circ} 30^{\prime} \mathrm{N}$ (LEZ/* ${ }^{*} 6 \mathrm{AN} 58$ ).

| Species: | Megrims <br> Lepidorhombus spp. | Zone: | 6; United Kingdom and international <br> waters of 5b; international waters of 12 <br> and 14 <br> (LEZ/56-14) |
| :--- | :--- | :--- | :--- |
| Spain | 550 | (1) | Analytical TAC <br> Article 8(2) of this Regulation applies |
| France | 2146 | (a) |  |
| Ireland | 627 | (a) |  |
| Union | 3323 | (2) |  |
| United Kingdom | 2258 | (1) |  |

TAC 5581
(a) Special condition: of which up to $25 \%$ may be fished in United Kingdom and Union waters of 2a and 4 (LEZ/*2AC4C).

2022 TAC for 6, EC waters of 5.b and International waters of 12 and 14 (lower) and TAC for 4 and 2.a (upper).

The uptake of the 2020 TAC for ICES Division 6 and EU waters of $5 . \mathrm{b}$ was $36.7 \%$. The small uptake was mainly due to poor utilisation of quota by France and the UK, managing only 5.8 and $37.4 \%$ respectively. In Area 4 and 2.a, uptake of the TAC was $78.4 \%$. The majority of available TAC ( $96.2 \%$ ) is allocated to the UK, who take $86.1 \%$ of it.

## Fishery in 2021

## Landings

Catches of megrim comprise two species, Lepidorhombus whiffiagonis and L. boscii. Information available to the Working Group indicates that L. boscii, are a negligible proportion of the Scottish and Irish megrim catch (Kunzlik et al., 1995; Anon, 2001).

Commercial catches are dominated by female megrim, typically $90 \%$ of the total catch. The InterCatch catch estimate is 3803 tonnes, and the ICES landings estimate for $6 . a$ and $4 . a$. is 3603 tonnes. The total ICES landings are well below the total TAC covering the fished areas of 4.a-6.a.

Official landings data for each country together with Working Group best estimates of landings from 6.a are shown in Table 10.2 and for 4.a in Table 10.3. To estimate ICES landings, we take InterCatch estimates and, if unavailable, we use official estimates. There are often minor differences between official data and InterCatch for most countries.

## Discards

Discard data were made available by Ireland, Scotland and France and total discards were estimated to be 200 tonnes or $5.6 \%$ by weight for the stock area in 2021. Total discard estimates have been reasonably consistent around $5-10 \%$ over the last nine years, although there have been some changes in rates within countries.

A linear decline in discards from 30 to 15\% over time between 1985 and 2012 is assumed in the stock assessment. From 2013 onwards discard data have taken from InterCatch, there is no deviation from the agreed stock annex.

## Catch

A breakdown of 2021 catch by main gear type in InterCatch is given below:

| Catch | Landings |  |  | Discards |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finfish trawls | Nephrops trawls | Other Gears | Finfish trawls | Nephrops trawls | Other Gears |
| 3803 tonnes | 98\% | <1\% | 1.6\% | 34\% | 66\% | <1\% |
|  |  | 3603 tonnes |  |  | 200 tonnes |  |

## Surveys

Indices from six fishery-independent surveys are used in the assessment. The surveys are outlined in Table 10.1 below and details can be viewed in the stock annex.

Table 10.1. Summary indices used for surplus production model.

| NUMBER | SURVEY | NATIONALITY | AREA | TIME-SERIES | DEPTH RANGE (M) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | $\begin{aligned} & \text { SCO-IBTS-Q3 } \\ & \text { (G2829) } \end{aligned}$ | Scotland | $4 . \mathrm{A}$ | 1987-PRESENT | <400 M |
| 2 | $\begin{aligned} & \text { SCO-IBTS-Q1 } \\ & (\mathrm{G} 1022) \end{aligned}$ | Scotland | $4 . \mathrm{A}$ | 1987-PRESENT | <400 м |
| 3 | $\begin{aligned} & \text { ScoGFS-WIBTS-Q1 } \\ & \text { (G1 179) } \end{aligned}$ | Scotland | 6.A | 1986-2010 | 40-400 |
| 4 | $\begin{aligned} & \text { ScoGFS-WIBTS-Q4 } \\ & \text { (G4299) } \end{aligned}$ | Scotland | 6.4 | 1986-2010 | 50-300 |
| 5 | SIAMISS-Q2 (G3745) | Scotland | 6.A*/4.A | 2005-PRESENT | 50-1050 |
| 6 | SIAMISS-Q2 (G1794) | Ireland | 6.A* | 2005-PRESENT | 50-850 |

Figures 10.1 to 10.5 present the megrim biomass maps for the SIAMISS and IBTS surveys. The SIAMISS bubble plots show and increasing abundance over time throughout the area over the time-series. Figures 10.2. (Sco-IBTS-Q3 (G2829) 4.a) and 13.3 (Sco-IBTS-Q1(G1022) 4.a) show the large increase in biomass over time in the northern North Sea. Biomass in the southern North Sea remains quite low.

Figures 10.4 (ScoGFS-WIBTS-Q1(G1179) 6.a) and 10.5 (ScoGFS-WIBTS-Q4(G4299) 6.a) also show an increase in biomass over the time-series and are shown until the survey design and ground gear changed in 2010. Data were truncated from the time-series going into the assessment.

### 10.2 Estimation of survey cpue indices

## Cpue trends of survey data

The data from the IBTS surveys exhibit a relatively large proportion of zeros, therefore the delta method of Stefánsson (1996) was used to generate indices. This method (delta-gamma model) comprises fitting two generalized linear models. The first model (binomial GLM) is used to obtain the proportion of non-zero tows, and is fit to the data coded as 1 or 0 , if the tow contained a positive or zero CPUE, respectively. The second model is fit to the positive only CPUE data using a gamma or lognormal GLM.

The biomass trend for the SIAMISS survey is shown in Figure 10.6. There is a weakly increasing trend over time with year effects evident in $6 . a$ in 2013 and 2017. The biomass trends for the four IBTS surveys are shown in Figure 10.7.

### 10.3 Stock assessment

The input data for the stock assessment are given in Table 10.4 this comprises of a time-series from all survey indices, and ICES catch estimates for this stock.

## 2022 Final run

The Pearson residual diagnostic plots for the final assessment are shown in Figure 10.8. The residuals for the two 6 .a surveys and the SIAMISS survey are fairly randomly dispersed around zero. A trend in the residuals is evident for the two 4 .a surveys, with increasing positive residuals in the last decade.

The prior and posterior distributions for the parameters in the final model fit, are shown in Figure 10.9. The priors are given in Table 10.5. The posterior distributions are similar to previous year's assessments. The posterior parameter estimates for the final assessment model are given in Table 10.6. These are similar to recent assessments.

Figure 10.10 shows the final model fits to the CPUE series and the estimates of total biomass and harvest ratio. The fits to the $6 . a$ and SIAMISS surveys are reasonable. The fits to the $4 . a$ surveys show that the model is not fitting well to those surveys in recent years. This issue needs to be examined further in the next benchmark.

The time-series of $B / B_{m s y}$ and $F / F_{m s y}$ landings and discards used in the final assessment are given in Table 10.7.

## Comparison with previous assessments

Figure 10.11 compares the final assessment with those conducted by WGCSE at previous meetings. The 2022 assessment assesses the biomass estimate to be stable at the 2021 levels; prior to being revised down in recent years. Estimates of fishing mortality continue on an upward trend. There are also some deviations in the historic estimates of F and Biomass around 2000. These are linked to the use of the 6.a surveys to derive the delta-gamma CPUEs truncated in 2010.

To evaluate evidence of possible bias in the assessment population metrics, a Mohn's Rho analysis resulted in values of -0.045 for $\mathrm{F}_{\mathrm{bar}}$ and 0.044 for biomass. ICES considers a value greater than 0.20 to be unacceptably high.

## State of the stock

The state of the stock has not changed since last year. Fishing mortality has been below Fmsy for almost the full time-series and has an overall declining trend since the late 1990s. Biomass has consistently been above MSY Btrigger and shows an increasing trend since 2005. The stock in 2022 is estimated to be 1.44 times $\mathrm{B}_{\text {MSY }}$ and the fishing mortality in 2021 is estimated to be have been $52 \%$ of FMSY.

### 10.4 Short-term projections

Short-term projections have been updated according to the method set out in the stock annex. The basis for the catch options is given in Table 10.8.

The management option table is given in Table 10.9. Fishing at $\mathrm{F}_{\text {MSY }}$ in 2023 is projected to result in total catches of 7200 t (landings of 6798 t and discards of 402 t ) and a Biomass of 1.32 times BMSY in 2024.

### 10.5 Biological reference points

## Precautionary approach reference points

Fmsy, Bmsy and the yield at MSY are all directly estimated in the model. It should be noted that these will vary when new survey and catch information is added. $\mathrm{B}_{\mathrm{pa}}$ and $\mathrm{B}_{\mathrm{lim}}$ are defined as $50 \% \mathrm{~B}_{\text {msy }}$ and $30 \% \mathrm{~B}_{\text {msy }}$ respectively. Flim is defined as $1.7 \mathrm{~F}_{\mathrm{MSY}}$ and is the F that drives the stock to $\mathrm{B}_{\mathrm{lim}}$ assuming $\mathrm{B}_{\mathrm{lim}}=30 \% \mathrm{~B}_{\mathrm{msy}}$. The derivation is given below:

```
P=rB(l-B/K)
The surplus productivity associated with Blim is:
P lim}=r\mp@subsup{B}{lim}{(1-B lim}/K
The corresponding F is:
F}\mp@subsup{F}{\mathrm{ lim }}{=r\mp@subsup{B}{lim}{lm}}(1-\mp@subsup{B}{\mathrm{ lim }}{}/K)/\mp@subsup{B}{\mathrm{ lim }}{}=r(1-\mp@subsup{B}{\mathrm{ lim }}{}/K
B lim}=0.3\mp@subsup{B}{MSY}{}=0.3K/
F}\mp@subsup{\mathrm{ lim }}{=r(1-0.3K/(2K))=r(1-0.3/2)=0.85r}{
F _ { M S Y } = r / 2 \text { , let x denote the proportionality between F} F _ { M S Y } \text { and F Flim}
xF (MSY }=\mp@subsup{F}{\mathrm{ lim}}{
x(r/2)=0.85r
x=2*0.85
x=1.7
```


## MSY reference points

In 2015 ICES provided precautionary $\mathrm{F}_{\mathrm{MSY}}$ ranges that are derived to deliver no more than a $5 \%$ reduction in long-term yield compared with MSY. Details of this analysis are given in WKMSYREF3 (ICES, 2015) and the derivations are given below.

|  | MSY Flower ${ }^{\text {b }}$ | $\mathrm{F}_{\text {MSY }}{ }^{\text {b }}$ | MSY $\mathrm{Fupper}^{\text {b }}{ }^{\text {b }}$ with AR | MSY ${ }^{\text {trigger }}$ |
| :---: | :---: | :---: | :---: | :---: |
| Megrim in divisions 4.a and 6.a | $0.39 \times r^{\text {d }}$ ) | $\mathrm{r} / 2 \mathrm{~d}$ ) | $\mathrm{r} / 2 \mathrm{~d})$ | K d) |

The stock has been fished below Fmsy for more than ten years, therefore, the WG considered it appropriate to set the MSY $B_{\text {trigger }}=B_{\text {MSY }}$ according to the ICES guidelines (ICES, 2017).

## Uncertainties and bias in assessment and forecast

The model estimates of B and F do have large uncertainty. Despite this, there is a low probability that SSB is below MSY $\mathrm{B}_{\text {Trigger }}$ and a high probability that F is below $\mathrm{F}_{\text {msY. }}$

The reference points are re-estimated within the assessment. The change between 2022 and 2021 reference points are consistent with previous years and results in a rescaling of relative stock status. However, in absolute terms, stock trends are consistent with those of previous years.

The biomass time-series from surveys has increasing uncertainty boundaries as the index increases. This results in uncertainty bounds in the model estimates; shows a contraction from the 2021 assessment.

Owing to incomplete discard data, historical discard rates (1985-2012) are assumed to have declined, from $30 \%$ at the beginning of the time-series, to an estimate of $15 \%$ in 2012 . The evaluation
of current stock status is robust to this assumption. Estimates since 2013 are based on observed discards.

## Recommendation for next benchmark

This stock was subject to an inter-benchmark in 2012 (IBP-MEG, 2012). Due to incomplete age data, particularly for 4.a, a Bayesian state-space surplus production model was chosen as the final assessment model. Subsequent update assessments have highlighted a problem fitting to the 4 .a surveys which needs to be examined in a future benchmark.

WGCSE recommends the following explorations:

- The SIAMISS survey should be merged into one continuous index. The length data for the index should also be examined.
- The ScoGFS-WIBTS-Q1/Q4 2011+: the ScoGFS-WIBTS-Q1/Q4 survey time-series should also be examined for re-introduction into the assessment as a new time-series. There may also be scope to integrate the IGFS.
- Available length and age-structured data should be compiled for this stock.
- Length or age-structured assessment models could be explored.

Once sufficient progress has been made on the points above, WGCSE will suggest a benchmark schedule.

## Management considerations

Megrim is a bycatch species in the mixed demersal trawl in divisions 6.a and 4.a. Management measures for other species have constrained the fishery and reduced effort and fishing mortality on megrim. The general increase in mesh size in 6 and 4 since 2010 has also benefited the stock.
The TAC in 6 has not been fully utilised. However, the uptake rate is country-specific, with some Member States reporting landings above their quota in the North Sea. Partial quota uptake by individual Member States may be linked to reduction in effort rather than reflective of a reduction in biomass. The TAC and assessment area are incompatible. There are two separate TAC areas covering ICES areas 6 and 4, whereas the assessment covers ICES divisions 6.a and 4.a combined. Due consideration of the inconsistency between management and assessment area is required when setting fishing opportunities for this stock and the separate 6.b Rockall stock. ICES (2013) have advised the EC that the TAC areas should be consistent with the assessment area and that ICES has no basis on how to split the catch advice so that it is consistent with the TAC areas.

### 10.6 References

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Table 10.2. Megrim in Subarea 6.a. Nominal catch ( $t$ ) of Megrim West of Scotland, as officially reported to ICES and WG best estimates of landings. The shaded cells show updates in official data compared with last year.

| $\begin{aligned} & \frac{1}{\pi} \\ & \end{aligned}$ | $\begin{aligned} & \frac{E}{工} \\ & \frac{B}{60} \\ & \infty \end{aligned}$ | $\begin{aligned} & \text { ※ } \\ & \text { 든 } \\ & \text { 끈 } \end{aligned}$ | $\begin{aligned} & \text { ס } \\ & \text { 들 } \\ & \underline{\underline{N}} \end{aligned}$ |  | $\begin{aligned} & \stackrel{\text { II }}{01} \\ & \text { in } \end{aligned}$ | $\begin{aligned} & \frac{y}{0} \\ & n_{n}^{n} \\ & \text { iod } \\ & \text { ud } \\ & \frac{1}{y} \end{aligned}$ | $\frac{\dot{\bar{y}}}{\dot{\sum}}$ |  | $\underset{ }{〕}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 0 | 398 | 317 | 0 | 91 | 25 |  | 1093 | - | 1924 | 2210 |
| 1991 | 1 | 455 | 260 | 0 | 48 | 167 |  | 1223 | - | 2154 | 2432 |
| 1992 | 0 | 504 | 317 | 0 | 25 | 392 |  | 887 | - | 2125 | 2549 |
| 1993 | 0 | 517 | 329 | 0 | 7 | 298 |  | 896 | - | 2047 | 2721 |
| 1994 | 1 | 408 | 304 | 0 | 1 | 327 |  | 866 | - | 1907 | 2693 |
| 1995 | 0 | 618 | 535 | 0 | 24 | 322 |  | 952 | - | 2451 | 3498 |
| 1996 | 0 | 462 | 460 | 0 | 22 | 156 |  | 944 | - | 2044 | 4054 |
| 1997 | 0 | 192 | 438 | 1 | 87 | 123 |  | 954 | - | 1795 | 3272 |
| 1998 | 0 | 172 | 433 | 0 | 111 | 65 |  | 841 | - | 1622 | 2705 |
| 1999 | 0 | 0 | 438 | 0 | 83 | 42 |  | 831 | - | 1394 | 2648 |
| 2000 | 0 | 135 | 417 | 0 | 98 | 20 |  | 754 | - | 1424 | 2247 |
| 2001 | 0 | 252 | 509 | 0 | 92 | 7 |  | 770 | - | 1630 | 2473 |
| 2002 | 0 | 79 | 280 | 0 | 89 | 14 |  | 643 | - | 1105 | 1828 |
| 2003 | 0 | 92 | 344 | 0 | 98 | 13 |  | 558 | - | 1105 | 1642 |
| 2004 | 0 | 50 | 278 | 0 | 45 | 17 |  | 469 | - | 859 | 1328 |
| 2005 | 0 | 48 | 156 | 0 | 69 | 10 |  | 269 | - | 552 | 561 |
| 2006 | 0 | 53 | 221 | 0 | 52 |  |  |  | 346 | 672 | 875 |
| 2007 | 0 | 104 | 191 | 0 | 5 |  |  |  | 667 | 967 | 1301 |
| 2008 | 0 | 92 | 172 | 0 | 149 |  |  |  | 874 | 1287 | 1545 |
| 2009 | 0 | 174 | 188 | 0 | 112 |  |  |  | 953 | 1427 | 1387 |
| 2010 | 0 | 271 | 318 | 0 | 288 |  |  |  | 822 | 1699 | 1698 |
| 2011 | 0 | 153 | 227 | 0 | 217 |  |  |  | 715 | 1312 | 1297 |
| 2012 | 0 | 140 | 214 | 0 | 142 |  |  |  | 590 | 1086 | 1132 |
| 2013 | 0 | 105 | 203 | 0 | 213 |  |  |  | 470 | 991 | 949 |
| 2014 | 0 | 126 | 246 | 0 | 57 |  |  |  | 465 | 894 | 948 |
| 2015 | 0 | 140 | 311 | 0 | 140 |  |  |  | 520 | 1110 | 1110 |


|  | $\begin{aligned} & \frac{E}{7} \\ & \frac{1}{600} \\ & \underset{\infty}{0} \end{aligned}$ |  |  |  | 듳 ஸ゙ |  |  | $\underset{ }{〕}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | 0 | 189 | 408 | 0 | 146 |  |  | 694 | 1437 | 1437 |
| 2017 | 0 | 132 | 336 | 0 | 313 |  |  | 579 | 1359 | 1359 |
| 2018 | 0 | 117 | 301 | 0 | 273 |  |  | 680 | 1370 | 1392 |
| 2019 | 0 | 122 | 271 | 0 | 368 |  |  | 844 | 1606 | 1611 |
| 2020* | 0 | 119 | 250 | 0 | 302 |  |  | 710 | 1381 | 1380 |
| 2021* | 0 | 123 | 378 | 0 | 335 |  |  | 633 | 1468 | 1464 |

* Preliminary official landings.

Table 10.3. Megrim in Subarea 4 and 2.a. Nominal catch ( $t$ ) of Megrim North Sea, as officially reported to ICES and WG best estimates of landings.

| Z <br>  <br>  <br> 0 | $\begin{aligned} & \frac{\xi}{3} \\ & \frac{1}{W_{0}} \\ & \infty \end{aligned}$ |  | 뿐 둔 |  |  | $\begin{aligned} & \underset{\sim}{\bar{C}} \\ & \underline{\pi} \\ & \underline{\underline{N}} \end{aligned}$ |  | $\begin{aligned} & \text { 㐅 } \\ & \text { 3 } \\ & 0 \\ & \text { 2 } \end{aligned}$ | $\begin{aligned} & \text { 드̃ } \\ & \text { in } \end{aligned}$ | $\begin{aligned} & \stackrel{\widetilde{0}}{0} \\ & \stackrel{N}{3} \end{aligned}$ |  |  |  |  | Ј |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 4 | 2 | - | - | 3 | - | 24 | - | - | - | 17 | - | - | 1126 | - | 1176 | 837 |
| 1991 | 3 | 1 | - | 6 | - | - | 28 | - | - | - | 9 | - | - | 1169 | - | 1216 | 878 |
| 1992 | 2 | 4 | 36 | 3 | - | - | 27 | - | - | - | 47 | - | - | 1372 | - | 1491 | 1025 |
| 1993 | 7 | 6 | 25 | 4 | - | - | 30 | - | - | - | 8 | - | - | 1736 | - | 1816 | 1081 |
| 1994 | 2 | 1 | 27 | 1 | - | - | 28 | - | - | - | 19 | - | - | 2000 | - | 2078 | 1207 |
| 1995 | 7 | 2 | 24 | 2 | - | - | 26 | - | - | - | 44 | - | - | 2193 | - | 2298 | 1172 |
| 1996 | 5 | 7 | 14 | 1 | - | - | 9 | - | - | - | 4 | - | - | 3221 | - | 3261 | 1199 |
| 1997 | 3 | 5 | 16 | 2 | - | - | 20 | - | - | - | 3 | - | - | 3091 | - | 3140 | 1584 |
| 1998 | 5 | 18 | 14 | 4 | - | - | 30 | - | - | - | 5 | - | - | 2628 | - | 2704 | 1548 |
| 1999 | 4 | 21 | . | 1 | - | - | 26 | - | - | - | 4 | - | - | 2121 | - | 2177 | 1111 |
| 2000 | 10 | 29 | 7 | 3 | - | - | 20 | - | - | - | 2 | - | - | 2044 | - | 2115 | 1247 |
| 2001 | 2 | 52 | 5 | 1 | - | - | 11 | - | - | - | 2 | - | - | 1854 | - | 1927 | 1098 |


| $\begin{aligned} & \text { Z } \\ & \vdots \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \varepsilon \frac{\varepsilon}{\bar{b}} \\ & \frac{0}{\boxed{0}} \\ & \hline \end{aligned}$ |  |  | $\begin{aligned} & \text { त } \\ & \stackrel{0}{0} \\ & \stackrel{y}{0} \\ & 00 \end{aligned}$ |  | $\begin{aligned} & \text { 들 } \\ & \underline{0} \\ & \underline{\underline{0}} \end{aligned}$ |  | $\begin{aligned} & \text { 㐅} \\ & \sum_{0}^{3} \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { 듳 } \\ & \text { in } \end{aligned}$ |  |  | $\begin{aligned} & \infty \\ & 0 \\ & \frac{0}{0} \\ & \frac{\pi}{00} \\ & \frac{\tilde{0}}{\sqrt{n}} \\ & \frac{1}{3} \\ & \frac{1}{3} \end{aligned}$ |  | $\begin{aligned} & \text { 들 } \\ & \stackrel{\pi}{4} \\ & \stackrel{0}{n} \\ & \stackrel{1}{j} \end{aligned}$ | $\underset{ }{\checkmark}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 5 | 8 | 6 | - | - | - | 9 | - | - | - | 3 | - | - | 1675 | - | 1706 | 975 |
| 2003 | 3 | 11 | 11 | 2 | - | 1 | 7 | <0.5 | - | - | 1 | - | - | 1235 | - | 1271 | 727 |
| 2004 | - | 7 | 9 | 2 | - | - | 11 | <0.5 | - | - | 1 | - | - | 1130 | - | 1160 | 739 |
| 2005 | - | 1 | 3 | 4 | - | - | 19 | <0.5 | - | - | 1 | - | - | 958 | - | 986 | n/a |
| 2006 | 0 | 3 | 4 | 1 |  | 0 | 6 | 1 | 0 | 0 |  |  |  |  | 1342 | 1357 | 1179 |
| 2007 | 0 | 11 | 18 | 4 |  | 0 | 1 | 1 | 0 | 0 |  |  |  |  | 1437 | 1472 | 1047 |
| 2008 | 0 | 31 | 20 | 1 |  | 0 | 1 | 4 | 0 | 0 |  |  |  |  | 1524 | 1581 | 1349 |
| 2009 | 0 | 54 | 9 | 0 |  | 0 | 0 | 6 | 0 | 0 |  |  |  |  | 1474 | 1543 | 1484 |
| 2010 | 0 | 22 | 1 | 0 |  | 0 | 1 | 2 | 0 | 0 |  |  |  |  | 1440 | 1466 | 1499 |
| 2011 | 0 | 23 | 10 | 3 |  | 0 | 0 | 1 | 0 | 0 |  |  |  |  | 1394 | 1431 | 1421 |
| 2012 | 0 | 35 | 5 | 3 |  | 0 | 0 | 1 | 0 | 0 |  |  |  |  | 1397 | 1441 | 1458 |
| 2013 | 0 | 48 | 7 | 3 |  | 0 | 0 | 17 | 0 | 0 |  |  |  |  | 1690 | 1765 | 1788 |
| 2014 | 0 | 35 | 7 | 1 |  | 0 | 0 | 12 | 0 | 0 |  |  |  |  | 1475 | 1530 | 1551 |


| $\begin{aligned} & \text { Z } \\ & \vdots \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \varepsilon \frac{\varepsilon}{\bar{b}} \\ & \frac{0}{\boxed{0}} \\ & \infty \end{aligned}$ |  | U 든 둔 | $\begin{aligned} & \text { त } \\ & \stackrel{0}{0} \\ & \stackrel{y}{0} \\ & 00 \end{aligned}$ |  | $\begin{aligned} & \text { 믈 } \\ & \stackrel{\pi}{0} \\ & \underline{\underline{0}} \end{aligned}$ |  | $\begin{aligned} & \text { 又 } \\ & \text { 3 } \\ & \text { 300 } \end{aligned}$ | $\begin{aligned} & \stackrel{\check{10}}{0} \\ & \stackrel{0}{n} \end{aligned}$ | $\begin{aligned} & \stackrel{\smile}{\otimes} \\ & \stackrel{\rightharpoonup}{0} \\ & \stackrel{y}{3} \end{aligned}$ |  |  |  |  | 当 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 0 | 26 | 1437 | 0 |  | 0 | 0 | 8 | 0 | 0 |  |  |  |  | 1175 | 1217 | 1230 |
| 2016 | 0 | 46 | 13 | 2 |  | 0 | 2 | 21 | 0 | 0 |  |  |  |  | 1278 | 1362 | 1361 |
| 2017 | 0 | 60 | 36 | 3 |  | 0 | < 0.5 | 29 | 0 | 0 |  |  |  |  | 1199 | 1327 | 1235 |
| 2018 | 0 | 61 | 67 | 1 |  | 0 | 1 | 34 | 0 | 0 |  |  |  |  | 1543 | 1706 | 1611 |
| 2019 | 0 | 63 | 103 | 4 |  | 0 | 1 | 46 | 0 | 0 |  |  |  |  | 1340 | 1557 | 1585 |
| 2020* | $<0.5$ | 40 | 80 | 3 |  | 0 | 4 | 48 | 0 | < 0.5 |  |  |  |  | 1768 | 1943 | 1935 |
| 2021* | 0 | 73 | 74 | 9 |  | 0 | 1 | 49 | 0 | $<0.5$ |  |  |  |  | 1944 | 2150 | 2139 |

* Preliminary official landings.

Table 10.4. Time-series of megrim survey indices in ICES Area 6.a and Division 4 as used in the surplus production model.

| year | sco.6.a.q1 | sco.6.a.q4 | sco.4.a.q1 | sco.4.a.q3 | monk.6.a | monk.4.a |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 2.587 | NA | NA | NA | NA | NA |
| 1986 | 1.688 | NA | 1.288 | NA | NA | NA |
| 1987 | 1.371 | NA | 1.525 | NA | NA | NA |
| 1988 | 2.009 | NA | 1.721 | NA | NA | NA |
| 1989 | 1.162 | NA | 1.345 | NA | NA | NA |
| 1990 | 1.073 | 1.589 | 0.771 | NA | NA | NA |
| 1991 | 0.793 | 1.274 | 0.509 | 0.331 | NA | NA |
| 1992 | 0.958 | 1.885 | 0.654 | 0.319 | NA | NA |
| 1993 | 1.013 | 2.058 | 1.106 | 0.306 | NA | NA |
| 1994 | 1.589 | 3.246 | 0.270 | 0.381 | NA | NA |
| 1995 | 1.556 | 1.863 | 0.000 | 0.391 | NA | NA |
| 1996 | 1.940 | 1.946 | 0.516 | 0.605 | NA | NA |
| 1997 | 1.100 | 1.081 | 0.433 | 0.430 | NA | NA |
| 1998 | 1.094 | 1.893 | 0.836 | 0.224 | NA | NA |
| 1999 | 1.322 | 1.360 | 1.005 | 0.237 | NA | NA |
| 2000 | 1.140 | 1.186 | 0.869 | 0.249 | NA | NA |
| 2001 | 0.998 | 0.968 | 0.297 | 0.092 | NA | NA |
| 2002 | 0.760 | 1.857 | 1.269 | 0.352 | NA | NA |
| 2003 | 1.272 | 1.205 | 0.519 | 0.316 | NA | NA |
| 2004 | 1.244 | 1.064 | 0.283 | 0.460 | NA | NA |
| 2005 | 0.690 | 1.013 | 0.590 | 0.809 | 1660.379 | 4753.223 |
| 2006 | 0.917 | 1.121 | 0.790 | 0.927 | 2688.942 | 3344.997 |
| 2007 | 0.907 | 1.199 | 0.868 | 1.389 | 3380.351 | 6347.544 |
| 2008 | 1.253 | 0.957 | 1.607 | 1.195 | 2467.080 | 7754.168 |
| 2009 | 1.573 | 1.397 | 1.949 | 1.100 | 3830.668 | 5946.946 |
| 2010 | 1.171 | NA | 1.769 | 1.726 | 3312.129 | 5394.946 |
| 2011 | NA | NA | 1.983 | 1.638 | 2501.990 | 4683.594 |


| year | sco.6.a.q1 | sco.6.a.q4 | sco.4.a.q1 | sco.4.a.q3 | monk.6.a | monk.4.a |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | NA | NA | 2.609 | 1.523 | 3450.807 | 4839.468 |
| 2013 | NA | NA | 2.669 | 1.477 | 6174.864 | 6460.015 |
| 2014 | NA | NA | 2.202 | 1.277 | 3033.072 | 11970.300 |
| 2015 | NA | NA | 3.014 | 1.297 | 2563.105 | 4986.899 |
| 2016 | NA | NA | 1.440 | 1.274 | 3027.648 | 8207.787 |
| 2017 | NA | NA | 1.830 | 1.013 | 6508.563 | 10238.937 |
| 2018 | NA | NA | 1.414 | 1.073 | 3364.165 | 7154.307 |
| 2019 | NA | NA | 0.657 | 0.963 | 2143.573 | 7982.271 |
| 2020 | NA | NA | 1.362 | 0.866 | NA | NA |
| 2021 | NA | NA | 1.154 | 0.736 | 3268.490 | 6897.872 |

Table 10.5. Lepidorhombus whiffiagonis in ICES areas 6.a and 4.a. Prior distributions on parameters.

| Parameter | Symbol | Prior distribution | Notes |
| :---: | :---: | :---: | :---: |
| Intrinsic rate of population growth | $r$ | Uniform(0.001,2.0) |  |
| Carrying capacity | K | $\text { Uniform }\left(\ln (\max (C)), \ln \left(10 \times \sum_{t=1985}^{2010} c_{t}\right)\right.$ | From the maximum catch to ten times the cumulative catch across all years assuming uniform distribution on the logarithmic scale |
| Catchabilities | $\log \left(q_{j}\right)$ | Uniform( $-11.0,0.0$ ) | Uniformly distributed on log-scale. See catchability sensitivity in Section 2.2.3.1 |
| Process error variance | $\frac{1}{\sigma_{u}^{2}}$ | Gamma $($ shape $=0.001$, rate $=0.001)$ | Gamma distributed on inverse variance (precision) scale |
| Measurement error variances | $\frac{1}{\sigma_{\varepsilon, j}^{2}}$ | Gamma $($ shape $=0.001$, rate $=0.001)$ | Gamma distributed on inverse variance (precision) scale |
| $\begin{aligned} & \text { Proportion of } K \text { in } \\ & 1985 \end{aligned}$ | $a$ | Uniform(0.01, 2.0) |  |

Table 10.6. Parameter estimates for final assessment outputs.

| Parameter | Esti- <br> mates <br> 2014 | Estimates 2015 | Esti- <br> mates <br> 2016 | Esti- <br> mates $2017$ | Estimates 2018 | Estimates 2019 | Estimates 2020 | Estimates 2021 | Esti- <br> mates <br> 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| r.hat | 0.55 | 0.51 | 0.51 | 0.51 | 0.47 | 0.50 | 0.51 | 0.52 | 0.52 |
| K.hat | 43134 | 47216 | 46840 | 42681 | 55129 | 44116 | 42625 | 41634 | 40573 |
| MSY | 5660 | 5612 | 5362 | 5072 | 5362 | 5123 | 5101 | 5020 | 4978 |
| $\mathrm{F}_{\text {MSY }}$ | 0.28 | 0.26 | 0.26 | 0.25 | 0.23 | 0.25 | 0.26 | 0.26 | 0.26 |
| $\mathrm{B}_{\text {MSY }}$ | 21567 | 23608 | 23420 | 21340 | 27565 | 22058 | 21313 | 20817 | 20287 |
| B | 4109 | 42416 | 42356 | 37610 | 38057 | 37062 | 32660 | 32408 | 31632 |
| F | 0.08 | 0.07 | 0.07 | 0.07 | 0.08 | 0.08 | 0.1 | 0.1 | 0.12 |
| $\mathrm{Blim}^{\text {m }}$ | 6470 | 7082 | 7026 | 6402 | 8269 | 6617 | 6394 | 6245 | 6086 |
| $\mathrm{B}_{\text {trig }}$ | 10783 | 11804 | 11710 | 10670 | 13782 | 11029 | 10656 | 10408 | 10143 |

Table 10.7. Time-series of $B / B_{\text {MSY }}$ and $F / F_{\text {MSY }}$ estimates and landings and discards in tonnes for the final assessment.

| Year | Low | $B / B_{M S Y}$ <br> Value | High | Landings | Discards* | Low | F/FMSY <br> Value | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 1.20 | 2.33 | 3.54 | 4499 | 1928 | 0.38 | 0.68 | 1.16 |
| 1986 | 1.08 | 1.64 | 2.25 | 2858 | 1193 | 0.33 | 0.53 | 0.77 |
| 1987 | 1.07 | 1.54 | 2.05 | 4614 | 1874 | 0.58 | 0.93 | 1.31 |
| 1988 | 1.02 | 1.43 | 2.04 | 5212 | 2061 | 0.65 | 1.15 | 1.66 |
| 1989 | 0.80 | 1.15 | 1.56 | 3451 | 1327 | 0.50 | 0.88 | 1.27 |
| 1990 | 0.74 | 1.05 | 1.43 | 3047 | 1140 | 0.47 | 0.82 | 1.22 |
| 1991 | 0.70 | 0.99 | 1.30 | 3310 | 1204 | 0.59 | 0.95 | 1.40 |
| 1992 | 0.72 | 1.04 | 1.38 | 3574 | 1263 | 0.58 | 0.98 | 1.46 |
| 1993 | 0.79 | 1.12 | 1.51 | 3802 | 1305 | 0.57 | 0.97 | 1.42 |
| 1994 | 0.81 | 1.21 | 1.70 | 3900 | 1300 | 0.53 | 0.93 | 1.32 |
| 1995 | 0.86 | 1.24 | 1.75 | 4670 | 1511 | 0.59 | 1.10 | 1.57 |
| 1996 | 0.82 | 1.19 | 1.71 | 5253 | 1649 | 0.68 | 1.31 | 1.91 |
| 1997 | 0.72 | 1.00 | 1.37 | 4856 | 1478 | 0.78 | 1.40 | 2.01 |
| 1998 | 0.66 | 0.95 | 1.37 | 4253 | 1254 | 0.67 | 1.26 | 1.86 |
| 1999 | 0.62 | 0.92 | 1.36 | 3759 | 1074 | 0.59 | 1.13 | 1.70 |
| 2000 | 0.57 | 0.86 | 1.24 | 3494 | 966 | 0.56 | 1.10 | 1.64 |
| 2001 | 0.55 | 0.80 | 1.11 | 3571 | 956 | 0.65 | 1.20 | 1.80 |
| 2002 | 0.55 | 0.82 | 1.18 | 2803 | 725 | 0.47 | 0.88 | 1.32 |
| 2003 | 0.58 | 0.86 | 1.21 | 2369 | 592 | 0.37 | 0.70 | 1.06 |
| 2004 | 0.59 | 0.87 | 1.22 | 2067 | 499 | 0.32 | 0.58 | 0.88 |
| 2005 | 0.63 | 0.89 | 1.16 | 1527 | 356 | 0.24 | 0.41 | 0.61 |
| 2006 | 0.72 | 1.01 | 1.31 | 2054 | 461 | 0.30 | 0.49 | 0.73 |
| 2007 | 0.82 | 1.15 | 1.47 | 2348 | 508 | 0.31 | 0.50 | 0.73 |
| 2008 | 0.91 | 1.28 | 1.66 | 2894 | 602 | 0.34 | 0.56 | 0.83 |
| 2009 | 1.01 | 1.40 | 1.84 | 2871 | 574 | 0.30 | 0.51 | 0.72 |
| 2010 | 1.03 | 1.41 | 1.85 | 3197 | 614 | 0.36 | 0.56 | 0.84 |
| 2011 | 1.03 | 1.45 | 1.91 | 3257 | 600 | 0.29 | 0.47 | 0.68 |


| Year | B/BMSY <br> Value |  | High | Landings | Discards* | F/FMSY <br> Value |  | High |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 1.16 | 1.58 | 2.08 | 2545 | 449 | 0.26 | 0.42 | 0.60 |
| 2013 | 1.27 | 1.77 | 2.43 | 2737 | 327 | 0.22 | 0.37 | 0.53 |
| 2014 | 1.33 | 1.79 | 2.36 | 2500 | 309 | 0.20 | 0.33 | 0.48 |
| 2015 | 1.21 | 1.67 | 2.17 | 2471 | 152 | 0.21 | 0.31 | 0.44 |
| 2016 | 1.28 | 1.75 | 2.26 | 2792 | 167 | 0.23 | 0.36 | 0.51 |
| 2017 | 1.37 | 1.88 | 2.62 | 2594 | 193 | 0.19 | 0.32 | 0.45 |
| 2018 | 1.25 | 1.69 | 2.16 | 3003 | 255 | 0.26 | 0.41 | 0.56 |
| 2019 | 1.12 | 1.55 | 1.96 | 3197 | 184 | 0.29 | 0.45 | 0.65 |
| 2020 | 1.13 | 1.56 | 2.07 | 3316 | 214 | 0.30 | 0.47 | 0.68 |
| 2021 | 1.14 | 1.53 | 1.99 | 3603 | 200 | 0.34 | 0.52 | 0.74 |
| 2022 | 0.95 | 1.44 | 2.01 |  |  |  |  |  |

* Discard estimates prior to 2013 are approximated, based on limited sampling information.

Table 10.8. Basis for the catch options.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| $\mathrm{F}_{2022} / \mathrm{F}_{\mathrm{MSY}}$ | 0.52 | Status quo: $\mathrm{F}_{\mathrm{sq}}=$ relative $\mathrm{F}(2021)$ |
| $\mathrm{B}_{2023} / \mathrm{B}_{\mathrm{MSY}}$ | 1.53 | Fishing at $\mathrm{F}_{\mathrm{sq}}$ |
| Catch (2022) | 3580 | Fishing at $\mathrm{F}_{\mathrm{sq}} ;$ in tonnes |
| Projected land- <br> ings (2022) | 3380 | Assuming average landings ratio (2019-2021); in tonnes |
| Projected dis- <br> cards (2022) | 200 | Assuming average discard ratio (2019-2021); in tonnes |

Table 10.9. The management option table.

| Basis | Total catch (2023) | Projected landings (2023) | Projected discards (2023) | Fishing mortality $\mathrm{F}_{\text {2023 }} / \mathrm{F}_{\mathrm{MSY}}$ | Stock size <br> $\mathrm{B}_{2024} / \mathrm{B}_{\mathrm{MSY}}$ | \% B change* | \% TAC change^ | \% advice change ${ }^{\wedge \wedge}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |
| MSY approach $=\mathrm{F}_{\text {MSY }}$ | 7200 | 6798 | 402 | 1 | 1.32 | -13.1 | -13.7 | -2.0 |
| EU MAP^^^A: $\mathrm{F}_{\text {MSY }}$ | 7200 | 6798 | 402 | 1 | 1.32 | -13.1 | -13.7 | -2.0 |
| $F=M A P \wedge \wedge \wedge ~ F_{\text {MSY lower }}$ | 5550 | 5240 | 310 | 0.76 | 1.41 | -7.1 | -33 | -24 |
| $F=M A P \wedge \wedge \wedge \wedge F_{\text {MSY upper }}$ | 7200 | 6798 | 402 | 1 | 1.32 | -13.1 | -13.7 | -2.0 |
| $F=0$ | 0 | 0 | 0 | 0 | 1.71 | 11.7 | -100 | -100 |
| $\mathrm{B}_{2024}=\mathrm{Bl}_{\text {lim }}$ | 26800 | 25302 | 1498 | 3.70 | 0.33 | -79 | 221 | 264 |
| $\mathrm{B}_{2024}=\mathrm{B}_{\mathrm{pa}}=\mathrm{MSY} \mathrm{B}_{\text {trigger }}$ | 13000 | 12273 | 727 | 1.80 | 1.01 | -33 | 56 | 77 |
| $\mathrm{B}_{2024}=\mathrm{B}_{2023}$ | 3500 | 3304 | 196 | 0.48 | 1.52 | 0 | -58 | -52 |

* Biomass 2024 relative to biomass 2023.
$\wedge$ Total catch in 2023 relative to TAC 2022 ( 8344 tonnes).
$\wedge \wedge$ Advice value for 2023 relative to the advice value for 2022 ( 7350 tonnes).
$\wedge \wedge \wedge$ EU multiannual plan (MAP) for the Western Waters (EU, 2019).


Figure 10.1. Maps of the northern continental shelf around the British Isles showing the biomass of megrim during the Scottish Irish Anglerfish and Megrim Industry Science Survey (SIAMISS) survey 2005-2021. There was no survey in 2020 due to COVID.


Figure 10.2. Sco-IBTS Q3 4.a 1991-present megrim biomass maps.


Figure 10.3. ScoIBTS Q1 4.a 1986-present megrim biomass maps.


Figure 10.4. ScoGFS-WIBTS Q1 6.a megrim biomass maps.


Figure 10.5. ScoGFS-WIBTS Q4 6.a megrim biomass maps.


Figure 10.6. Megrim biomass estimates in ICES divisions 4, 6.a and 6.b from Scottish Irish Anglerfish and Megrim Industry Science Survey (SIAMISS) survey with 95\%cls.


Figure 10.7. Megrim CPUE estimates in ICES Division 6.a Q1 top left panel and 6.a Q4.


Figure 10.8. Pearson residuals for the six survey indices.


Figure 10.9. Prior (red line) and posterior distributions (black line) for the parameters in the model.


Figure 10.10. Time-series of catch and model estimates of total biomass and exploitation rate (median values are shown as solid lines and $95 \%$ confidence intervals shown as broken lines). The model fits to the various CPUE series is also shown (observations dots, median fit solid line and 95\% confidence intervals shown as broken lines).


Figure 10.11. Comparison with previous assessments.


Figure 10.12. Kobe plot of stock status.

## 11 Megrim (Lepidorhombus spp.) in Division 6.b (Rockall)

## Type of assessment in 2022

This stock was benchmarked in 2021 (ICES, 2021) and, as a result, the stock was changed from category 3 to category 2. The assessment, which is now based on Surplus Production in Continuous Time (SPiCT, Pedersen and Berg, 2017), includes revised assumptions and model priors. Reference points were also revised. These changes have resulted in a more reliable assessment and the methodology is appropriate to determine stock status and a short-term catch forecast.

## ICES advice applicable to 2023

ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, catches in 2023 should be no more than 1022 tonnes.

## General

## Stock description and management units

Megrim stock structure is uncertain. Data collected during an EC study contract (98/096) on the 'Distribution and biology of anglerfish and megrim in the waters to the west of Scotland,' showed significantly different growth parameters and significant population structure difference between megrim sampled in $6 . a$ and $6 . b$ (Gordon, 2001). Spawning fish occur in both areas but whether these populations are reproductively isolated is not clear. WKFLAT (2011) concluded that megrim in $6 . b$ should continue to be considered as a separate stock until further information is available.


TAC/Management area
[ Assessment area

Management area (red box) and assessment area (blue hatched area).

TAC regulations for 2021 and 2022 are given below:

| Species: | Megrims <br> Lepidorhombus spp. |  | Zone:6; United Kingdom and international waters of <br> 5b; international waters of 12 and 14 <br> (LEZ/56-14) |
| :--- | :--- | :--- | :--- | :--- |
| Spain | 526 | (1) | Analytical TAC |
| France | 2053 | (1) | Article 8(2) of this Regulation applies |
| Ireland | 600 | (1) |  |
| Union | 3179 | (1) |  |
| United | 2046 | (1) |  |
| Kingdom | 5225 |  |  |
| TAC |  |  |  |

${ }^{(1)} \quad$ Special condition: of which up to $25 \%$ may be fished in: United Kingdom and Union waters of 2a and 4 (LEZ/*2AC4C).

2022:

| Species: | Megrims <br> Lepidorhombus spp. |  |  | Zone: | 6; United Kingdom and international waters of 5 b ; international waters of 12 and 14 <br> (LEZ/56-14) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Spain |  | 550 | (1) | Analytical TAC Article 8(2) of this Regulation applies |  |
| France |  | 2146 | ${ }^{(2)}$ |  |  |
| Ireland |  | 627 | (1) |  |  |
| Union |  | 3323 | (2) |  |  |
| United Kingdom |  | 2258 | (1) |  |  |
| TAC |  | 5581 |  |  |  |

## Fishery in 2021

Ireland had the highest catches in 2021 followed by the UK and Spain (Table 14.1). The majority of the landings and catches are from otter trawlers.

| Catch | Landings |  | Discards |
| :---: | :---: | :---: | :---: |
| 631 tonnes | Otter trawls $85 \% \quad$ Other gears 15\% | Otter trawls 100\% |  |
|  | 566 tonnes | 65 tonnes |  |

## Landings

Official landings data for each country together with Working Group best estimates of landings from 6.b are shown in Table 14.1 and Figure 14.1.

Catches of megrim comprise two species, Lepidorhombus whiffiagonis and L. boscii. Information available to the Working Group indicates that L. boscii, are a negligible proportion of the Scottish and Irish megrim catch (Kunzlik et al., 1995; Anon, 2001). It is not clear to the WG whether landings of other countries are accurately partitioned by megrim species. Megrim are caught in association with anglerfish by some fleets and are area-misreported along with anglerfish. However, it is unknown whether misreporting from Division 6.6 is an issue.

## Discards

Discard data for 2021 were available for Ireland, UK and Spain in InterCatch. Total discard estimates were available from 2005-2021. To estimate catches prior to 2005, for the SPiCT analysis; a catch over landing ratio of 1.2 was used (derived from that observed ratio between 2017-2020). In 2021, discards represented approximately $10 \%$ of catch; increasing to 65 from 59 tonnes (Table 14.1 and Figure 14.1).

## Surveys

In 2005, Scotland initiated a new industry-science partnership survey to provide an absolute abundance estimate for anglerfish. Sixteen years of survey data are available and these cover the main distribution of the anglerfish fishery. The survey is also considered to have greater spatial coverage for megrim, and as such was recommended by WKAGME (2008) as the main source of data of megrim relative biomass, for all megrim stocks in the Northern Shelf.

The survey index for $6 . b$ is shown in Figure 14.2. Due to technical reasons the survey was unable to sample in Division 6.b, so the stock size is unknown for 2022. This value is also absent in 2020 due to the absence of the SCO-IV-VIAMISS-Q2 [G1794] survey cancellation due to Covid. Sensitivity trials showed the assessment to be robust to the missing data and it was decided by the group to use the updated assessment despite the missing input data. Details and outputs of the sensitivity trials are included below in the uncertainties section of the report.

The available data shows the stock abundance to have been stable since 2012 prior to which it displayed a largely increasing abundance and biomass trend since 2005. The area-stratified survey provides a minimum estimate of absolute biomass; survey catches are raised based on swept area and weighted by area. The survey assumes that all megrim in the trawl path are retained e.g. $q=1$. Assuming full retention is overly optimistic, therefore the minimum estimate of stock biomass was provided.

## Historical stock development

Prior to the benchmark in 2021, the stock was a category 3 stock that utilised a SPiCT assessment and the ratio of the mean of the last two SCO-IV-VIAMISS-Q2 index values.

## Final Assessment

Following on from the benchmark (ICES, 2021), the final assessment utilised a SPiCT model utilising the recommendations and developed settings. The catch data is shown in Figure 14.1 and combined the landings and discard estimates. The abundance index from the SCO-IV-VIAMISSQ2 survey is shown in Figure 14.2. Following on from the sensitivity and robustness testing at the benchmark the following prior settings were applied:

- Surplus production curve fixed $(\mathrm{n}=2)$
- Intrinsic growth rate (r) 0.39 - modelled from FishLife
- An initial biomass depletion prior of 0.5
- Intermediate year catch - average of last 3 years' catch

The output of the model can be seen in Figure 14.3. The residuals are good (Figure 14.4) and the retrospective plots for the assessment show good agreement with all the peels (Figure 14.5). Final parameter estimates from the SPiCT run are given in Table 14.2.

## State of the stock

The summary plots can be seen in Figure 14.3 and they show fishing pressure on the stock is below $\mathrm{F}_{\text {MSY }}$ and biomass is above MSY $\mathrm{B}_{\text {trigger }}$ and $\mathrm{Blim}_{\text {lim }}$.

## Short-term projections

Short term projections were conducted using a 2022 catch that was the average of the preceding 3 years, and the assumptions are shown below:

| Variable | Value | Notes |
| :--- | :---: | :--- |
| F (2022)/FMSY | 0.62 | F corresponding to status quo catch |
| B (2023)/BMSY | 1.31 | Short term forecast (STF) with status quo catch |
| Catch (2022) | 759 | Status quo catch (average 2019-2021) ; in tonnes |

Four management scenarios were explored and the catch and relative reference points estimated for 2023. Adopting the MSY approach (using the $35^{\text {th }}$ percentile of predicted catch under $\mathrm{F}=\mathrm{F}$ MSY) gave an estimated catch of 1022 tonnes, a F/Fmš of 0.91 and a B/BMSY of 1.26.

| Basis | Total catch <br> $(2023)$ | $\mathrm{F}_{2023} / \mathrm{F}_{\mathrm{MSY}}$ | $\mathrm{B}_{2024} / \mathrm{B}_{\mathrm{MSY}}$ | \% B change |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |
| MSY approach $\left(35^{\text {th }}\right.$ percentile of predicted <br> catch distribution under $\left.\mathrm{F}=\mathrm{F}_{\mathrm{MSY}}\right)$ | 1022 | 0.91 | 1.26 | -3.67 |  |
| $\mathrm{~F}_{\mathrm{MSY}}$ |  |  |  |  |  |
| $\mathrm{F}_{2021}$ | 1116 | 1.00 | 1.24 | -5.17 |  |
| F $=0$ | 793 | 0.69 | 1.31 | -0.06 |  |

## MSY reference points

The MSY reference points are calculated based on the relative reference points estimated by the SPiCT model, so will change when the assessment is updated. The reference points are calculated as:

| Framework | Reference point | Value | Technical basis |
| :---: | :---: | :---: | :---: |
| MSY approach | MSY Btrigger | 0.5 * | Relative value ( $\mathrm{B} / \mathrm{BMsY}$ ) from the SPiCT assessment model. Bmsy is estimated directly from the SPiCT model and changes when the assessment is updated. |
|  | FMSY | 1 * | Relative value ( $\mathrm{F} / \mathrm{Fmsy}$ ) from the SPiCT assessment model. FmSY is estimated directly from the SPiCT model and changes when the assessment is updated. |
| Precautionary approach | Blim | $0.3 \times$ BMSY | Relative value (equilibrium yield at this biomass is $50 \%$ of MSY). |
|  | Flim | $1.7 \times$ FMSY | Relative value (the F that drives the stock to $\mathrm{B}_{\mathrm{lim}}$ ). |

## Uncertainties and bias in assessment and forecast

Due to the missing 2022 and 2020 survey data sensitivity analyses were conducted to determine the impact on the assessment. Comparisons between the 2021 assessment and the updated 2022
were conducted, and showed that the $\mathrm{r}, \mathrm{K}$ and q parameters (shown in the table below) estimated were very similar, with less than $3 \%$ change for each. There were some differences in the estimates states in the assessments, with lower fishing mortality and higher biomass states in 2022. This would be expected due to the catches being reduced by $31 \%$ in 2021.

| Parameter echry | 2021 Assessment | 2022 Assessment | Percent.Change |
| :---: | :---: | :---: | :---: |
| r | 0.379 | 0.384 | 1.379 |
| K | 9421.813 | 9177.065 | -2.598 |
| q | 0.615 | 0.618 | 0.440 |
| B_2021.38 | 5903.066 | 5938.855 | 0.606 |
| F_2021.38 | 0.156 | 0.105 | -32.834 |
| B_2021.38/Bmsy | 1.259 | 1.301 | 3.321 |
| F_2021.38/Fmsy | 0.829 | 0.549 | -33.732 |

Figure 14.6 compares the time series of the relative statuses as estimated by last year's assessment and the updated assessment. There is very good agreement up until 2022, at which point the assessment would be expected to diverge due to the large decrease in catches (31\%) in 2021. The updated assessment produces a slight increase in advice from $2021(<3 \%)$, based on these lower catches, and lower intermediate year catch assumptions.

To evaluate the assessments robustness to missing data, scenarios was recreating using the complete time series as used in 2019. Three assessments were compared:

- Full time series up to 2019
- Missing index in 2019
- Missing index in 2017 and 2019

The parameters and the state of the stock are very similar in all three scenarios. Comparing the full scenario and missing two index values found the r and k parameter estimates to be less than $1 \%$ different. The survey catchability parameter differed more ( $11 \%$ ) but that might be expected to be more poorly estimated as the index time series becomes significantly shortened. The fishing mortality and biomass estimates differed approximately $5 \%$ between the assessments. Overall the time series of relative states show good agreement (Figure 14.7).

| Paramater | Full | Missing Last | Missing Last and -2 |
| :---: | :---: | :---: | :---: |
| r | 0.374 | 0.375 | 0.377 |
| K | 9725.512 | 9525.894 | 9641.080 |
| q | 0.609 | 0.602 | 0.679 |
| B_2019.38 | 6180.054 | 6109.830 | 5904.964 |
| F-2019.38 | 0.156 | 0.159 | 0.165 |
| B_2019.38/Bmsy | 1.278 | 1.291 | 1.234 |
| F_2019.38/Fmsy | 0.836 | 0.853 | 0.880 |

The catch advice from these different assessment was also in good agreement, ranging from 1036 tonnes in the full assessment down to 992 with both missing data points.
Both these analyses provide reasonable reassurance that the current assessment is robust to missing survey data to the current extent. It also shows that the catch advice is unlikely to have been greatly impacted. There is some concern due to the large decrease in catch in 2021 when there is no available estimate of abundance to verify the health of the stock. It appears that this is partially due to decreased fishing effort in the area, although some decrease in lpue is also noted.
Currently the assessment uses the SCO-IV-VIAMISS-Q2 survey to estimate biomass. It should be noted that the survey was specifically designed to catch angler fish. While this is not an issue when the biomass index is presented in the relative context, in the case of megrim; the raised
biomass calculation is based on full retention of megrim in the haul. The estimates are therefore considered as the minimum.

## Recommendation for next Benchmark

This stock was subject to benchmark in 2021.

## Management considerations

The TAC in 6 has not been fully utilised; the uptake rate is country-specific; partial quota by individual Member States may be an artefact of reduction in effort rather than reflective of a reduction in biomass. The TAC and assessment area are incompatible.

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Table 14．1 Megrim in Division 6．b．History of catch and landings；official landings presented by country and ICES estimated catch．All weights are in tonnes．

| デ ત્兀 |  | $\begin{aligned} & \text { む్ } \\ & \text { む } \\ & \text { む̀ } \end{aligned}$ | $\begin{aligned} & \text { जु } \\ & \text { की } \end{aligned}$ |  | $\begin{aligned} & \mathscr{0} \\ & \frac{0}{n} \\ & 3 \\ & 2 \\ & 0 \\ & \frac{\pi}{80} \\ & \frac{1}{4} \\ & \frac{1}{5} \end{aligned}$ |  | $\stackrel{V}{\square}$ |  |  | 0 0 0 0 0 0 0 0 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | － | 240 | 587 | 14 | － | 204 | － | 1045 | 1045 |  |
| 1992 | － | 139 | 683 | 53 | － | 198 | － | 1073 | 1073 |  |
| 1993 | － | 128 | 594 | 56 | － | 147 | － | 925 | 925 |  |
| 1994 | － | 176 | 574 | 38 | － | 258 | － | 1046 | 1046 |  |
| 1995 | － | 117 | 520 | 27 | － | 152 | － | 816 | 816 |  |
| 1996 | － | 124 | 515 | 92 | － | 112 | － | 843 | 843 |  |
| 1997 | － | 141 | 628 | 76 | － | 164 | － | 1009 | 1009 |  |
| 1998 | － | 218 | 549 | 116 | － | 208 | － | 1091 | 1091 |  |
| 1999 | － | 127 | 404 | 57 | － | 278 | － | 866 | 866 |  |
| 2000 | 4 | 167 | 427 | 57 | － | 309 | － | 964 | 964 |  |
| 2001 | $<0.5$ | 176 | 370 | 42 | － | 236 | － | 824 | 824 |  |
| 2002 | ＜0．5 | 87 | 120 | 41 | － | 207 | － | 455 | 455 |  |
| 2003 | － | 83 | 93 | 74 | － | 382 | － | 632 | 632 |  |
| 2004 | － | 43 | 71 | 42 | － | 372 | － | 528 | 528 |  |
| 2005 | － | 68 | 88 | 19 | － | 207 | － | 382 | 382 | 87 |
| 2006 | － | 95 | 59 | － | － | － | 190 | 344 | 344 | 75 |
| 2007 | － | 87 | 19 | － | － | － | 188 | 106 | 106 | 22 |
| 2008 | － | 68 | 84 | － | － | － | 142 | 294 | 294 | 59 |
| 2009 | － | 48 | 46 | － | － | － | 165 | 226 | 226 | 44 |
| 2010 | － | 47 | 41 | － | － | － | 93 | 139 | 139 | 26 |
| 2011 | － | 72 | 28 | － | － | － | 69 | 155 | 155 | 7 |
| 2012 | － | 120 | 61 | － | － | － | 89 | 224 | 224 | 21 |
| 2013 | － | 181 | － | － | － | － | 58 | 278 | 278 | 15 |
| 2014 | － | 230 | 73 | － | － | － | 95 | 343 | 343 | 15 |
| 2015 | － | 256 | 190 | － | － | － | 130 | 453 | 453 | 85 |
| 2016 | － | 272 | 69 | － | － | － | 116 | 405 | 405 | 145 |
| 2017 | － | 358 | 215 | － | － | － | 180 | 586 | 586 | 233 |
| 2018 | － | 438 | 61 | － | － | － | 263 | 762 | 764 | 203 |
| 2019 | 25 | 76 † | 94 | － | － | － | 229 | 791 | 757 | 34 |
| 2020＊ | 41 | 467 | 112 | － | － | － | 246 | 866 | 861 | 59 |
| 2021＊ | 1 | 293 | 71 |  |  |  | 212 | 577 | 566 | 65 |
|  |  |  |  |  |  |  |  |  |  |  |

＊Landing values are preliminary．
† Incomplete／missing as a result of part of the data being unavailable under data confidentiality clauses．

Catches


Figure 14.1. Lez.27.6b ICES estimated landings and discards. Discard data are only available since 2005; values prior to that are assumed to be $20 \%$ of landings based on the observed ratio from 2017 to 2020.


Figure 14.2. Survey data for lez.6b from SIAMISS (SCO-IV-VI-AMISS-Q2 [G1794])


Figure 14.3. SPiCT model output for lez.27.6b. Top right: observed and fitted catch with 95 ci. Centre left: Biomass relative to $B_{\text {MSY }}$. Centre: $F$ relative to $F_{\text {MSY }}$. Corresponding MSY quantities are shown in each plot as horizontal lines ( $0.5 \mathrm{~B}_{\mathrm{MSY}}$ in the case of the relative biomass plot). Centre right Kobe plot of stock trajectory.


Figure 14.4. SPiCT model residual output for lez.27.6b

Number of retrospective years


Figure 14.5. SPiCT model retrospectives for lez.27.6b

## Table14.2. SPICT results for Lez.27.6b.

| Model parameter estimates w 95\% CI |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
|  | estimate | cilow | ciupp | log.est |
| alpha | 5.9604221 | 0.7450337 | $4.768460 \mathrm{e}+01$ | 1.7851413 |
| beta | 0.3342888 | 0.0982301 | $1.137625 \mathrm{e}+00$ | -1.0957499 |
| r | 0.3838002 | 0.2388906 | $6.166110 \mathrm{e}-01$ | -0.9576333 |
| rc | 0.3838002 | 0.2388906 | $6.166110 \mathrm{e}-01$ | -0.9576333 |
| rold | 0.3838002 | 0.2388906 | $6.166110 \mathrm{e}-01$ | -0.9576333 |
| m | 880.5398435 | 619.4560299 | $1.251663 \mathrm{e}+03$ | 6.7805352 |
| K | 9177.0656892 | 4863.6232775 | $1.731601 \mathrm{e}+04$ | 9.1244628 |
| q | 0.6176870 | 0.2677990 | $1.424715 \mathrm{e}+00$ | -0.4817734 |
| sdb | 0.0551011 | 0.0071406 | $4.251910 \mathrm{e}-01$ | -2.8985853 |
| sdf | 0.2790944 | 0.1823418 | $4.271848 \mathrm{e}-01$ | -1.2762053 |
| sdi | 0.3284259 | 0.2253436 | $4.786626 \mathrm{e}-01$ | -1.1134440 |
| sdc | 0.0932981 | 0.0363117 | $2.397173 \mathrm{e}-01$ | -2.3719552 |
| pp | 0.9615441 | 0.7651621 | $9.948154 \mathrm{e}-01$ | 3.2190290 |
| robfac | 10.4473688 | 2.3561872 | $6.681155 \mathrm{e}+01$ | 2.2457363 |

Deterministic reference points (Drp)

|  | estimate | cilow | ciupp | log.est |
| :--- | ---: | ---: | ---: | ---: |
| Bmsyd | 4588.5328446 | 2431.8116387 | 8658.0034944 | 8.431316 |
| Fmsyd | 0.1919001 | 0.1194453 | 0.3083055 | -1.650780 |
| MSYd | 880.5398435 | 619.4560299 | 1251.6633606 | 6.780535 |

Stochastic reference points (Srp)
estimate cilow ciupp log.est rel.diff.Drp
Bmsys $4566.32662732423 .04639278605 .42287998 .426464-0.004863037$
Fmsys $0.1911496 \quad 0.1189586 \quad 0.3071504-1.654699-0.003926166$
$\begin{array}{llllllll}\text { MSYs } & 872.8348518 & 617.1145708 & 1234.5206457 & 6.771746 & -0.008827548\end{array}$

| States w 95\% CI (inp\$msytype: s) |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
|  | estimate | cilow | ciupp | log.est |
| B_2021.94 | 5938.8541864 | 2620.3230175 | $1.346017 \mathrm{e}+04$ | 8.6892715 |
| F_2021.94 | 0.1050109 | 0.0420162 | $2.624536 \mathrm{e}-01$ | -2.2536909 |
| B_2021.94/Bmsy | 1.3005759 | 0.8310666 | $2.035333 \mathrm{e}+00$ | 0.2628071 |
| F_2021.94/Fmsy | 0.5493651 | 0.2379788 | $1.268188 \mathrm{e}+00$ | -0.5989920 |

Predictions w 95\% CI (inp\$msytype: s)

|  | prediction | cilow | ciupp | log.est |
| :--- | ---: | ---: | ---: | ---: |
| B_2024.00 | 6223.0147558 | 2763.9332488 | $1.401116 \mathrm{e}+04$ | 8.7360098 |
| F_2024.00 | 0.1050113 | 0.0314164 | $3.510072 \mathrm{e}-01$ | -2.2536875 |
| B_2024.00/Bmsy | 1.3628054 | 0.8951825 | $2.074704 \mathrm{e}+00$ | 0.3095454 |
| F_2024.00/Fmsy | 0.5493670 | 0.1743656 | $1.730870 \mathrm{e}+00$ | -0.5989886 |
| Catch_2023.00 | 647.0016731 | 324.4560955 | $1.290194 \mathrm{e}+03$ | 6.4723489 |
| E(B_inf) | 6583.3951800 | NA | NA | 8.7923059 |



Figure 14.6. Comparison of relative time series from the SPiCT assessments for lez.27.6b in 2021 and 2022


Figure 14.7. Sensitivity analyses comparing time series from the SPiCT assessments for lez.27.6b in with missing survey data points

# 12 Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 11 (West of Scotland, North Minch) 

Nephrops stocks have previously been identified by WGNEPH on the basis of population distribution, and defined as separate Functional Units. The Functional Units (FU) in ICES Division 6.a (of which there are three) are defined by the groupings of ICES statistical rectangles given in Table 12.1 and illustrated in Figure 12.1. The functional unit is the level at which the WG collates fishery data (quantities landed and discarded, fishing effort and length distributions) and at which it performs assessments.

## Type of assessment in 2022

The assessment of North Minch Nephrops in 2022 is based on a combination of examining trends in fishery indicators and abundance estimated by underwater TV survey, both of which comprise an extensive data series for this FU. The assessment follows the process defined by the benchmark WG (WKNEPH 2009 and WKNEPH 2013) and is conducted annually according to standards set out by the Manual for the Nephrops Underwater TV Surveys (Dobby H., et al, 2021). Further details on the assessment and catch options are provided in the stock annex.

## ICES advice applicable to 2021

'ICES advises that when the EU multiannual plan (MAP) for Western waters and adjacent waters is applied, catches in 2021 that correspond to the F ranges in the MAP are between 3075 tonnes and 3953 tonnes. The entire range is considered precautionary when applying the ICES advice rule.

To ensure that the stock in Functional Unit 11 is exploited sustainably, management should be implemented at the functional unit level.'

## ICES advice applicable to 2022

'ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2018-2020, catches in 2022 should be no more than 3853 tonnes.

To ensure that the stock in Functional Unit 11 is exploited sustainably, management should be implemented at the functional unit level.

ICES notes the existence of a management plan, developed and adopted by one of the relevant management authorities for Subarea 6. ICES considers this plan to be precautionary when implemented at the FU level.'

### 12.1 General

Nominal landings as reported to ICES for divisions $6 . \mathrm{a}$ and $6 . \mathrm{b}$ are presented in Table 12.1.1. Total official landings from Division 6.a were 9764.4 tonnes in 2021, mostly reported by the UK
with only 42 tonnes reported from Ireland. Table 12.1.2 and Figure 12.1.1 shows WG estimates of landings in Division 6.a broken down by FU. Nephrops landings are also made from outside the functional units, from statistical rectangles where small pockets of suitable sediment exist, although these are generally small amounts. In 2021, 237 tonnes of landings were reported from outside the FUs which is lower than the long-term average (Table 12.1.2). The main areas of activity outside FUs are the Stanton Bank (to the west of the South Minch) and areas of suitable sediment along the shelf edge and slope to the west of the Hebrides. There are no functional units in Division 6.b and only very small quantities of Nephrops are landed (Table 12.1.1(b)).

## Stock description and management units

The North Minch (FU 11) is located at the northern end of the west coast of Scotland (Figure 12.1). Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Within the North Minch functional unit these substrates are distributed according to prevailing hydrographic and bathymetric conditions. The area is characterised by numerous islands of varying size and sea lochs, which occur along the mainland coast. These topographical features create a diverse habitat with complex hydrography and a patchy distribution of soft sediments. Results from work on mapping the spatial extent of Nephrops habitat in the North Minch sea lochs indicate that the muddy habitat in these areas is only a very small proportion of the total Nephrops grounds in the North Minch (WKNEPH 2013).

## Management applicable to 2021 and 2022

The management unit is Subarea 6 and EU and international waters of 5.b. The TAC for this area is 12065 tonnes in 2022, down from 15294 tonnes in 2021.

Since 2016, fisheries catching Nephrops in Division 6.a have been covered by the EU landing obligation (EU, 2015a). Creel fisheries are exempt from the landing obligation due to high survivability of discards. Dermersal trawlers using a codend between $80 \mathrm{~mm}-110 \mathrm{~mm}$ and within 12 miles of shore are also exempt from the landings obligation.

## Ecosystem aspects

Details of the ecosystem aspects for this functional unit are provided in the stock annex if available.

## Fishery description

Information on developments in the fishery was provided by Marine Scotland compliance officers.

In 2021, the fishery was generally described as fairly good. It was reported that COVID-19 did not have much impact on the local fishery, with market remaining stable for most of the year.
In recent years the fishery starts steadily, with a good yield in the summer fishery from May to August. The fishery then tails off in the Autumn. This is said to be a seasonal occurrence rather than being caused by bad weather. The majority of the Western Isles trawl fleet has tended to relocate to the east coast and to the fishing grounds in the Firth of Forth/Eyemouth/Shields for the winter months in recent years. Trawl activity in the winter months is generally at a relatively low level.

Activity in the Nephrops trawl sector was up in 2021 owing to the relaxation of COVID-19 pandemic rules. However, the creel sector remained fairly stable.

The largest part of the North Minch fleet is still based at Stornoway, numbering approximately 75 vessels in 2021. The majority of the Stornoway vessels (52) are below 10 m in length.

The fleet were targeting the same areas in the North Minch as previous years. The notable changes were that the trawl fleet stayed in the West coast when in previous years they would go to the East coast from September onward. The trawl fleet also lost most of their summer fishing due to the COVID-19 pandemic.

Very few vessels came from outside to fish in the area and activity in the area overall has been reduced in 2021.

Since 2009, vessels have been required to fit 120 mm square meshed panels, in accordance with the west coast emergency measures (Council Reg. (EU) 43/2009). Large SMPs ( 200 mm ) are also widely used in the North Minch and have been mandatory for all TR2 vessels with power >112 kW fishing under the Scottish Conservation Credits scheme.
Further general information on the fishery can be found in the stock annex.

### 12.2 Data available

## InterCatch

Data for 2021 were successfully uploaded into InterCatch prior to the 2022 WG meeting. Uploaded data were worked up in InterCatch to generate 2021 raised international length-frequency distributions. Allocation schemes for any unsampled fleets are described in the stock annex. Data exploration in InterCatch has previously shown that outputs of raised data were very close to those generated by the previous method applied internally with differences being $<0.1 \%$. As such, InterCatch length-frequency outputs have been used in the stock assessment since 2012.

The COVID-19 pandemic resulted in a reduced sampling effort of commercial catches for FU11 in 2021, as was also the case in 2020. Discard sample data for FU11 were only available for Quarter three and four, and so, InterCatch estimates of discard rates for Quarters one and two in InterCatch were based on samples collected in Quarter three and four. Following download of data from InterCatch, alternate methods of 2021 discard estimation were thus considered. It was agreed at WGCSE that estimates of discard rates and associated size distributions for 2021 would be based on an averaging of discard samples across all quarters for which data are available between 2017 and 2021. Minimum and maximum discard rates over the same period were also examined to gain an appreciation of the plausible range of discarding that might have occur in 2021. Assessment estimates affected by changes in discard rates are annotated hereafter to reflect this; i.e. " $x(y / z)$ ", where $x$ is the estimate based on the average discard rate between 2017 and $2021, y$ is based on minimum discard rate, and $z$ on maximum discard rate over the same period.

## Commercial catch

Official catch statistics (landings) reported to ICES are shown in Tables 12.1.1(a) and 12.1.1(b); these relate to the whole of 6 .a of which the North Minch is a part. Landings by gear category for FU11 provided by country have been reported since 1981 and are presented in Table 12.2.1. Landings from this fishery are usually only reported from Scotland; between 2012 and 2014 two tonnes of Nephrops landings were reported by Ireland and values between one and three tonnes have been reported since 2017. Total reported Scottish landings in 2021 were 2073.1 tonnes, consisting of 1547 tonnes landed by trawlers targeting Nephrops ( $\sim 75 \%$ ), 472 tonnes landed by creel
vessels ( $\sim 23 \%$ ) and 53 tonnes by other trawlers. In 2021, a small amount of Nephrops below minimum size (BMS) was also reported (1.1 tonnes).

## Effort data

In 2015, WGCSE agreed that effort should be reported in kW days as this is likely to be more informative about changes in the actual fleet effort. Reported effort by Scottish trawlers targeting Nephrops (Métiers: OTB_CRU - Bottom Otter Trawls Targeting Crustaceans and OTT_CRU Multirig Otter Trawls Targeting Crustaceans) has shown a decreasing trend since 2000 (Figure 12.2.1) but in 2012 the effort increased by $20 \%$ due to the influx of vessels from the North Sea during the first quarter of the year. Since then, effort has declined although there was a small increase in 2016. The decline in effort observed in recent years continued until 2020. The observed increase in effort for 2021 may be attributed to the relaxation of the COVID-19 pandemic rules. Note that the year range in effort time-series (2000-2021) does not match the more extensive year range available for landings, due to a lack of confidence in the reliability of older effort data in the Marine Scotland Science database. The effort is also slightly inconsistent with the landings data because effort is provided for TR2 vessels only, while the 'Nephrops trawl' landings also includes landings by large mesh trawlers targeting Nephrops.

## Sampling levels

Length compositions of landings and discards are obtained during market and on-board observer sampling respectively. These sampling levels are shown in Table 12.2.2. Owing to the relaxation of COVID-19 pandemic rules, which disrupted both the fishing industry and government sampling programmes, sampling effort in 2021 was slightly higher compared to 2020, although still lower than recent years. Length compositions for the creel fishery are available for landings only as the small numbers of discards survive well and are not considered to be removed from the population.

## Length compositions

Figure 12.2.2 shows a series of annual length-frequency distributions for the period 2000 to 2021. Catch (removals) length compositions are shown for each sex along with the mean length for both. In both sexes the mean sizes fluctuate over time and has generally remained stable since 2012. This parameter might be expected to reduce in size if overexploitation were taking place. In 2021, the mean size of males increased while that of the female slight decreased when compared to 2020. s.

## Sex ratio

Males consistently make the largest contribution to the landings, although the proportion of males does vary between years (Figure 12.2.3(a)). This is likely due to the varying seasonal pattern in the fishery and associated relative catchability (due to different burrow emergence behaviour) of male and female Nephrops. Males are available throughout the year and the fishery is prosecuted in all quarters (although effort is usually reduced during the winter months when the weather is poor). Females are mainly taken in the summer when they emerge after egg hatching. The seasonal change in proportion of males to females is evident in Figure 12.2.3(b). In 2021 the normal temporal trend in sex ratios was observed where males dominate in quarters one and four but the ratio is more even (or often female dominated) in quarters two and three.

## Mean weights

The mean weight in the landings (trawls and creels combined) shows substantial interannual variation (Figure 12.2.4 and Table 12.2.3) decreasing between 2010 and 2012, followed by an increase in 2013-2015 and a decrease again in 2016 and stable in 2017 followed by an increasing trend onward. Given the relatively larger size of creel caught Nephrops (compared to trawl) the proportion of creel landings has a substantial effect on overall size composition. The increases in mean weight to 2010 and 2020 (and also size, Figure 12.2.2) in particular were due to a higher proportion of creel landings. Figure 12.2 .5 shows the mean weight by sample and gear type over the period 2011-2021. There is no obvious trend in North Minch trawl-caught mean weights for males and females, however, a decrease in the mean weight of creel caught males is still obvious. The mean weight in the landings has a significant impact on the catch forecast. Due to the high interannual variability in mean weights it was considered more appropriate to use a full timeseries average, from 1999 (first year with creel and trawl length distributions combined) until 2021 for producing the catch options.

## Discarding

Discarding of undersized and unwanted Nephrops occurs in this fishery, and quarterly discard sampling has been conducted on the Scottish Nephrops trawler fleet since 1990. Discard rates fluctuate in this FU and averaged $\sim 5.8 \%$ by number in the last three years (Table 12.2.4). In 2021, the discard rate increased to $6.2 \%$ by number (from $5.7 \%$ in 2020).

It is likely that some Nephrops survive the discarding process. An estimate of 25\% (Charuau et al., 1982; Sangster et al., 1997; Wileman et al., 1999) survival is assumed for this FU in order to calculate removals (landings + dead discards) from the population. The discard survival rate for creel caught Nephrops has been shown to be high (ICES, 2013) and a value of $100 \%$ is used. The discard rate (adjusted for survival) which will be used in the provision of landings options for 2023 is $5.8 \%$ based on a three-year average of 2019-2021.

## Abundance indices from UWTV surveys

The Underwater TV surveys are available for this stock since 1994 (missing surveys in 1995 and 1997). The stock area for this FU was updated in 2013 to $2908 \mathrm{~km}^{2}$ (see stock annex for further details). In 2022, due to the COVID-19 pandemic, the UWTV survey was carried out with a reduced scientific staffing, necessitating a reduced sampling schedule in some areas. UWTV survey in 2022 sampled $72 \%$ of the planned stations in relative to 2021 ( $100 \%$ of planned stations). While unquantified, the $28 \%$ reduction in the number of sampled stations is considered to have minimal impact on the quality of the abundance estimate.

A total of 36 valid TV stations were used in the final survey analysis (Table 12.2.5). Table 12.2.6 shows the basic analysis for the most recent TV survey conducted in FU11. At the 2012 SGNEPS meeting (ICES, 2012) it was decided that a CV (relative standard error) of $<20 \%$ was an acceptable precision level for UWTV survey estimates of abundance. The CV for the most recent TV survey was $13.2 \%$, lower than the precision level agreed (Table 12.2.6).

Figure 12.2 .6 shows the distribution of stations in recent TV surveys (2016-2022), with the size of the symbols reflecting the Nephrops burrow density. Table 12.2.5 and Figure 12.2.7 show the timeseries estimated abundance for the TV surveys, with $95 \%$ confidence intervals on annual estimates.

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009; ICES, 2013). A number of potential biases were highlighted including
those due to edge effects, species burrow misidentification and burrow occupancy. The cumulative relative to absolute conversion factor estimated for FU11 was 1.33 meaning that the TV survey is likely to overestimate Nephrops abundance by $33 \%$.

### 12.3 Assessment

## Comparison with previous assessments

The assessment follows the same procedure as last year and is based on a combination of examining trends in fishery indicators and underwater TV abundance estimates. Landings predictions are derived by applying a harvest rate to the UWTV survey estimate of abundance and assuming a length composition derived from recent fishery data (including data from both trawl and creel fisheries).

## State of the stock

The assessment summary is provided in Table 12.2.4. The underwater TV survey is presented as the best available information on the North Minch Nephrops stock. The surveys provide a fish-ery-independent estimate of Nephrops abundance. At present, it is not possible to extract any length or age-structure information from the survey and therefore it only provides information on abundance over the area of the survey.

TV survey estimated stock abundance in 2022 was 1346 million individuals, a 3.2\% decrease from the 2021 estimate. The stock is still well above the MSY B trigger value of 541 million, or the rounded value of 540 million individuals used in the provision of advice (Figure 12.2.7).

The calculated harvest rate in 2021 (dead removals/TV abundance $=4.6 \%$ ) is below the Fmsy proxy for this stock (the value associated with high long-term yield and low risk depletion) of $10.8 \%$.

### 12.4 Catch option table

Landings predictions and catch options at various harvest rates (based on principles established at WKNEPH (ICES, 2009)), including a selection of those equivalent to the per-recruit reference points, are made on the basis of the 2022 UWTV survey conducted in August (although normally in June). These were presented in September 2022 for the provision of advice.

The table below shows the agreed inputs to the catch options table.

| Input | Data | 2022 assessment |
| :--- | :---: | :---: |
| Survey abundance (millions) | UWTV 2022 | 1346 |
| Mean weight in projected landings (g) | $1999-2021$ | 26.54 |
| Mean weight in projected discards (g) | average 2019-2021 | $1999-2021$ |
| Dead projected discards | Proportion by number (assumed) | $5.8 \%^{*}$ |
| Discards survival rate | average 2019-2021 | $25 \%^{* *}$ |
| Dead discard rate |  | $4.4 \%$ |

Due to the high interannual variability in mean weights it was considered more appropriate to use a full time-series average, from 1999 (first year with creel and trawl length distributions combined) until 2021 for producing the catch options.
*Based on mean discard rate (2017-2021) allocated to Quarters 1 - 2 of 2021; estimates of $8.4 \%$ and $4.0 \%$ were derived based on the maximum and minimum observed discard rates, respectively, for the same period
** Discard survival in the creel fishery is assumed to be $100 \%$, as outlined in the stock annex.

### 12.5 Reference points

New reference point Fmsy were derived for this stock at WKMSYRef4 (ICES, 2016). This was updated on the basis of an average of estimated FMSY proxy harvest rates over a period of years, this corresponds more closely to the methodology for finfish. In cases where there is a clear trend in the values a five-year average was chosen. Similarly, the five-year average of the F at 95\% of the YPR obtained at the FmSy proxy reference point was proposed as the FmSy lower bound and the five-year average of the F above Fmax that leads to YPR of $95 \%$ of the maximum as the upper bound. Using an average value also has the advantage of reducing the effect of any unusually high or low estimates of the Fmsy proxy, which occasionally appear. For this stock, the FmSy proxy has been revised from $10.9 \%$ to $10.8 \%$.

WKFMSYRef4 did not update the MSY $B_{\text {trigger }}$ except for rounding to tens of millions. MSY Btrigger has been defined as the lowest stock size from which the abundance has increased (ICES, 2013) and is calculated as 541 million individuals and rounded to 540 million for use as MSY Btrigger in the advice. Full details are contained in the stock annex.

These reference points should remain under review by WGCSE and may be revised, should improve data become available.

Table 12.2.4 and Figure 12.5.1 show the harvest rates for FU11. From 2006-2009 there was a sustained period of high (above F msy proxy) harvest rates followed by two years of low harvest rates of around $6-7 \%$. A sudden increase was observed in 2012, following this, the harvest rate declined and has remained below the Fmsy proxy. Harvest rate historical low of $3.1 \%$ was recorded in 2020, with a slight increase to $4.6 \%$ in 2021 (still well below Fmsy proxy). It is likely that prior to 2006, the estimated harvest rates may not be representative due to underreporting of landings.

### 12.6 Management strategies

Scotland has recently established a network of regional Inshore Fisheries Groups (rIFGs), nonstatutory bodies that aim to improve the management of Scotland's inshore fisheries out to six nautical miles, and to give commercial inshore fishermen a strong voice in wider marine management developments. The rIFGs will contribute to regional policies and initiatives relating to management and conservation of inshore fisheries, including impacts on the marine environment and the maintenance of sustainable fishing communities and measures designed to better conserve and sustainably exploit stocks of shellfish and sea fish (including salmon) in their local waters. Although no IFG proposals specific to the management of Nephrops fisheries have yet been adopted, some of the IFG management plans for the Scottish West Coast include spatial management of Nephrops fisheries and the introduction of creel limits.

On the 8th of February 2016, phase 1 of the fisheries management measures for inshore MPAs in Scottish waters came into force (SG, 2016). These measures relate to both NCMPA (Marine (Scotland) Act and the UK Marine and Coastal Access Act) and SACs (EC Habitats Directives - Council Directive 92/43/EEC) both of which have the aim of conserving biological diversity in Scottish
waters and contribute to Scotland's MPA network (SG, 2017a). Although not specific to the management of the Nephrops fishery, they will influence spatial patterns of fishing for Nephrops where controls on the two main gear types, demersal trawls and creels are implemented on Nephrops habitat. Within the North Minch functional unit, two MPAs are covered by fisheries management measures. Specifically, the Wester Ross NCMPA where fishing activity is banned for demersal gears for vessels over 500 kW in power and banned in certain areas for vessels below 500 kW . North of the main Nephrops ground is the Loch Laxford SAC where demersal trawling is banned (SG, 2016). The areas of the SAC and NCMPA relative to the estimated Nephrops habitat within the North Minch functional unit are displayed in Figure 12.6.1.

### 12.7 Quality of assessment and forecast

The length and sex composition of the landings data is considered to be well-sampled. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in this fishery since 1990, and is considered to represent the fishery adequately. The reduced sampling effort in 2021 just like in 2020 meant that discard sample data were only available for Quarter 3 and 4, and it was agreed at WGCSE that estimates of discard rates and size distributions for Quarters 1 and 2 of 2021 would be adequately approximated for the purpose of forecasting by averaging of discard samples across all available Quarters between 2017 and 2021. The landings length compositions from 1999 onwards, are derived from both creel and trawl samples. The creel fishery which accounted for an increasing proportion of landings by $31 \%$ in 2020 has decrease in 2021 to $22.8 \%$. This part of the fishery exhibits a length distribution composed of larger animals.

There were concerns over the accuracy of historical landings and effort data prior to 2006 when Buyers and Sellers legislation was introduced and the reliability began to improve. Because of this, the final assessment adopted is independent of historical landings data. Harvest rates since 2006 are also considered more reliable due to more accurate landings data reported under this legislation. Incorporation of creel length compositions (since the 2010 WG ) has also improved estimates of harvest rates. Underwater TV surveys have been conducted for this stock since 1994, with a continual annual series available since 1998. The number of valid stations in the survey has remained relatively stable throughout the time period. Confidence intervals around the abundance estimates are relatively small for this functional unit. In the provision of catch options based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. A three-year average (i.e. 2019-2021 for the 2022 assessment) of discard rates (adjusted to account for some survival of discarded animals) has been used in the calculation of catch options.
The cumulative absolute conversion factor estimates for FU11 are largely based on expert opinion (see stock annex). The precision of these bias corrections cannot yet be characterised. The landings derived in the forecast (catch options table) are sensitive to the input dead discard rate and mean weights in landings, and this introduces uncertainties in the catch forecasts. Precision estimates are needed for these forecast inputs.

The stock area was revised in 2013 (ICES, 2013) using integrated VMS-logbook data to more accurately estimate the spatial extent of Nephrops catches. Two other factors however, have the potential to increase the fished area further. Firstly, the inclusion of vessels smaller than 15 m would likely increase the fished area in some of the inshore locations and secondly, it is known that most of the sea lochs have areas of mud substrate and are typically fished by creel boats. In recent years, a number of TV surveys have taken place in the major North Minch sea lochs in an attempt to improve estimates of the ground area and Nephrops abundance. Work presented at the WKNEPH 2013 (ICES, 2013) showed that the total area of the sea lochs is $105 \mathrm{~km}^{2}$, which is considerably smaller than the offshore VMS area estimated to be $2908 \mathrm{~km}^{2}$. Therefore, it is
unlikely that the exclusion of these inshore areas from the survey have an impact in the mean densities and overall abundance of Nephrops in the North Minch.

### 12.8 Recommendation for next benchmark

This stock was last benchmarked in 2013 (ICES, 2013). WGCSE will keep the stock under close review and recommend a future benchmark as required.

### 12.9 Management considerations

The WG, ACOM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level and management at the functional unit level could provide the controls to ensure that catch opportunities and effort were compatible and in line with the scale of the resource.

Creel fishing takes place in this area but overall effort by this fleet in terms of creel numbers is not known, and measures to control numbers are not in place. There is a need to ensure that the combined effort from all forms of fishing is taken into account when managing this stock.

There is a bycatch of other species in the area of the North Minch and STECF estimates that discards of whiting and haddock are high in 6.a generally. It is important that efforts are made to ensure that unwanted bycatch is kept to a minimum in this fishery. Efforts to reduce discards and unwanted bycatches of cod include the implementation of large square meshed panels (SMPs) of 120 mm under the west coast emergency measures, and SMPs of 200 mm which were introduced under the Scottish Conservation Credits scheme.

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Table 12.1. Nephrops functional units and descriptions by statistical rectangle.

| Functional Unit | Stock | Division | ICES Rectangles |
| :--- | :--- | :--- | :--- |
| 11 | North Minch | $6 . \mathrm{a}$ | $44-46$ E3-E4 |
| 12 | South Minch | $6 . \mathrm{a}$ | $41-43$ E2-E4 |
| 13 | Clyde | $6 . \mathrm{a}$ | $39-40$ E4-E5 |

Table 12.1.1(a). Nominal landings (tonnes) of Nephrops in Division 6.a, 1980-2021, as officially reported to ICES.

|  | France | Ireland | Spain | UK-(Engl+Wales+N.Irl) | UK- Scotland | UK | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 5 | 1 | - | - | 7422 | - | 7428 |
| 1981 | 5 | 26 | - | - | 9519 | - | 9550 |
| 1982 | 1 | 1 | - | 1 | 9000 | - | 9003 |
| 1983 | 1 | 1 | - | 11 | 10706 | - | 10719 |
| 1984 | 3 | 6 | - | 12 | 11778 | - | 11799 |
| 1985 | 1 | 1 | 28 | 9 | 12449 | - | 12488 |
| 1986 | 8 | 20 | 5 | 13 | 11283 | - | 11329 |
| 1987 | 6 | 128 | 11 | 15 | 11203 | - | 11363 |
| 1988 | 1 | 11 | 7 | 62 | 12649 | - | 12730 |
| 1989 | - | 9 | 2 | 25 | 10949 | - | 10985 |
| 1990 | - | 10 | 4 | 35 | 10042 | - | 10091 |
| 1991 | - | 1 | - | 37 | 10458 | - | 10496 |
| 1992 | - | 10 | - | 56 | 10783 | - | 10849 |
| 1993 | - | 7 | - | 191 | 11178 | - | 11376 |
| 1994 | 3 | 6 | - | 290 | 11047 | - | 11346 |
| 1995 | 4 | 9 | 3 | 346 | 12527 | - | 12889 |
| 1996 | - | 8 | 1 | 176 | 10929 | - | 11114 |
| 1997 | - | 5 | 15 | 133 | 11104 | - | 11257 |
| 1998 | - | 25 | 18 | 202 | 10949 | - | 11194 |
| 1999 | - | 136 | 40 | 256 | 11078 | - | 11510 |
| 2000 | 1 | 130 | 69 | 137 | 10667 | - | 11004 |
| 2001 | 9 | 115 | 30 | 139 | 10568 | - | 10861 |
| 2002 | - | 117 | 18 | 152 | 10225 | - | 10512 |
| 2003 | - | 145 | 12 | 81 | 10450 | - | 10688 |
| 2004 | - | 150 | 6 | 267 | 9941 | - | 10364 |
| 2005 | - | 153 | 17 | 153 | 7616 | - | 7939 |
| 2006 | - | 133 | 1 | 255 | 13419 | - | 13808 |


|  | France | Ireland | Spain | UK-(Engl+Wales+N.Irl) | UK- Scotland | UK | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | - | 155 | - | 2088 | 14120 | - | 16363 |
| 2008 | - | 56 | 1 | 419 | 14795 | - | 15271 |
| 2009 | - | 53 | - | 1226 | 11462 | - | 12741 |
| 2010 | - | 45 | 1 | 1962 | 10250 | - | 12258 |
| 2011 | - | 38 | - | 2517 | 10419 | - | 12974 |
| 2012 | - | 28 | - | 2502 | 11807 | - | 14337 |
| 2013* | - | 5 | - | - | - | 12866 | 12871 |
| 2014 | - | 51 | - | - | - | 12760 | 12811 |
| 2015 | - | 75 | - | - | - | 11653 | 11728 |
| 2016 | - | 107 | 0 | - | - | 14600 | 14707 |
| 2017 | - | 114 | - | - | - | 11442 | 11557 |
| 2018 | - | 65 | 0 | - | - | 8849 | 8914 |
| 2019 | - | 92 | - | - | - | 9 018* | 9110 |
| 2020 | - | 71 | - | 538 | 6334 | 6872 | 6943 |
| 2021 | - | 42 | - | 984 | 8738.4 | 9722.4 | 9764.4 |

* Includes 8.6 t landings reported by Isle of Man.

Table 12.1.1(b). Nominal landings (tonnes) of Nephrops in Division 6.b, 1980-2021 as officially reported to ICES. There are no Functional Units in ICES Division 6.b but occasional small landings are made.

|  | France | Germany | Ireland | Spain | UK-(Engl+Wales+N.Irl) | UK- Scotland | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | - | - | - | - | - | - | 0 |
| 1981 | - | - | - | - | - | - | 0 |
| 1982 | - | - | - | - | - | - | 0 |
| 1983 | - | - | - | - | - | - | 0 |
| 1984 | - | - | - | - | - | - | 0 |
| 1985 | - | - | - | - | - | - | 0 |
| 1986 | - | - | - | 8 | - | - | 8 |
| 1987 | - | - | - | 18 | 11 | - | 29 |
| 1988 | - | - | - | 27 | 4 | - | 31 |
| 1989 | - | - | - | 14 | - | - | 14 |
| 1990 | - | - | - | 10 | 1 | - | 11 |
| 1991 | - | - | - | 30 | - | - | 30 |
| 1992 | - | - | - | 2 | 4 | 1 | 7 |
| 1993 | - | - | - | 2 | 6 | 9 | 17 |
| 1994 | - | - | - | 5 | 16 | 5 | 26 |
| 1995 | 1 | - | - | 2 | 26 | 1 | 30 |
| 1996 | - | 6 | - | 5 | 65 | 5 | 81 |
| 1997 | - | - | 1 | 3 | 88 | 23 | 115 |
| 1998 | - | - | 1 | 6 | 46 | 7 | 60 |
| 1999 | - | - | - | 5 | 2 | 5 | 12 |
| 2000 | 2 | - | 8 | 3 | 4 | 4 | 21 |
| 2001 | 1 | - | 1 | 14 | 2 | 7 | 25 |
| 2002 | 1 | - | - | 7 | 3 | 7 | 18 |
| 2003 | - | - | 1 | 5 | 6 | 18 | 30 |
| 2004 | - | - | - | 2 | 7 | 13 | 22 |
| 2005 | 3 | - | 1 | 1 | 5 | 7 | 17 |
| 2006 | - | - | - | - | 1 | 3 | 4 |


|  | France | Germany | Ireland | Spain | UK-(Engl+Wales+N.Irl) | UK- Scotland | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | - | - | - | 2 | 3 | - | 5 |
| 2008 | - | - | - | - | - | - | 0 |
| 2009 | - | - | - | - | - | - | 0 |
| 2010 | - | - | - | - | - | - | 0 |
| 2011 | - | - | - | - | - | - | 0 |
| 2012 | - | - | - | - | - | - | 0 |
| 2013 | - | - | - | - | - | - | 0 |
| 2014 | - | - | - | - | - | - | 0 |
| 2015 | - | - | - | - | - | - | 0 |
| 2016 | - | - | - | - | - | 0 | 0 |
| 2017 | - | - | - | - | - | 2 | 2 |
| 2018 | - | - | - | - | - | 0 | 0 |
| 2019 | - | - | 0 | - | - | - | 0 |
| 2020 | - | - | 0.5 | - | - | - | - |
| 2021 | - | - | 0.02 | - | - | - | 0 |

Table 12.1.2. Nephrops, Total Nephrops landings (tonnes) by Functional Unit plus Other rectangles, 1981-2021.

| Year | FU11 | FU12 | FU13 | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 2861 | 3652 | 2968 | 39 | 9520 |
| 1982 | 2799 | 3552 | 2620 | 27 | 8998 |
| 1983 | 3197 | 3413 | 4076 | 34 | 10720 |
| 1984 | 4143 | 4300 | 3310 | 36 | 11789 |
| 1985 | 4060 | 4008 | 4286 | 104 | 12458 |
| 1986 | 3381 | 3484 | 4341 | 89 | 11295 |
| 1987 | 4084 | 3892 | 3009 | 257 | 11242 |
| 1988 | 4035 | 4473 | 3664 | 529 | 12701 |
| 1989 | 3205 | 4745 | 2812 | 212 | 10974 |
| 1990 | 2546 | 4430 | 2909 | 182 | 10067 |
| 1991 | 2793 | 4442 | 3038 | 255 | 10528 |
| 1992 | 3559 | 4237 | 2803 | 248 | 10847 |
| 1993 | 3193 | 4458 | 3343 | 344 | 11338 |
| 1994 | 3614 | 4414 | 2630 | 441 | 11099 |
| 1995 | 3655 | 4682 | 3987 | 460 | 12784 |
| 1996 | 2872 | 3995 | 4057 | 239 | 11163 |
| 1997 | 3046 | 4344 | 3621 | 243 | 11254 |
| 1998 | 2441 | 3730 | 4841 | 157 | 11169 |
| 1999 | 3257 | 4052 | 3752 | 438 | 11499 |
| 2000 | 3247 | 3953 | 3417 | 421 | 11038 |
| 2001 | 3259 | 3991 | 3182 | 420 | 10852 |
| 2002 | 3440 | 3305 | 3384 | 397 | 10526 |
| 2003 | 3269 | 3879 | 3173 | 433 | 10754 |
| 2004 | 3082 | 3869 | 2973 | 403 | 10327 |
| 2005 | 2949 | 3848 | 3395 | 254 | 10446 |
| 2006 | 4166 | 4633 | 4780 | 241 | 13820 |
| 2007 | 3978 | 5471 | 6660 | 420 | 16529 |
| 2008 | 3799 | 5356 | 5923 | 128 | 15206 |
| 2009 | 3496 | 4285 | 4779 | 185 | 12745 |


| Year | FU11 | FU12 | FU13 | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 2413 | 3846 | 5843 | 569 | 12671 |
| 2011 | 2697 | 3702 | 6432 | 219 | 13050 |
| 2012 | 3542 | 3989 | 6687 | 435 | 14653 |
| 2013 | 3413 | 3776 | 5435 | 234 | 12858 |
| 2014 | 3257 | 3179 | 6207 | 53 | 12696 |
| 2015 | 3002 | 3400 | 5147 | 309 | 11858 |
| 2016 | 3529.4* | 4402 | 6447 | 236 | 14614.4 |
| 2017 | 2491 | 3757 | 5403 | 250 | 11901 |
| 2018 | 1956 | 2540 | 4143 | 160 | 8799 |
| 2019 | 1979 | 2220 | 4683 | 173 | 9055 |
| 2020 | 1331 | 1976 | 3636 | 151 | 7094 |
| 2021 | 2073.1 | 2696.3 | 4995 | 237 | 10001.31 |

*Includes below minimum size landed discards of 0.4 t .

Table 12.2.1. Nephrops, North Minch (FU11), Nominal Landings of Nephrops, 1981-2021.

| UK Scotland |  |  |  |  |  | Other United Kingdom and Ireland | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | Nephrops trawl | other | creel | Below Minimum Size | Subtotal |  |  |
| 1981 | 2320 | 171 | 370 | 0 | 2861 | 0 | 2861 |
| 1982 | 2323 | 105 | 371 | 0 | 2799 | 0 | 2799 |
| 1983 | 2784 | 96 | 317 | 0 | 3197 | 0 | 3197 |
| 1984 | 3449 | 160 | 534 | 0 | 4143 | 0 | 4143 |
| 1985 | 3235 | 117 | 708 | 0 | 4060 | 0 | 4060 |
| 1986 | 2641 | 203 | 537 | 0 | 3381 | 0 | 3381 |
| 1987 | 3459 | 143 | 482 | 0 | 4084 | 0 | 4084 |
| 1988 | 3450 | 148 | 437 | 0 | 4035 | 0 | 4035 |
| 1989 | 2603 | 112 | 490 | 0 | 3205 | 0 | 3205 |
| 1990 | 1941 | 134 | 471 | 0 | 2546 | 0 | 2546 |
| 1991 | 2229 | 126 | 438 | 0 | 2793 | 0 | 2793 |
| 1992 | 2978 | 149 | 432 | 0 | 3559 | 0 | 3559 |
| 1993 | 2699 | 86 | 408 | 0 | 3193 | 0 | 3193 |
| 1994 | 2916 | 245 | 453 | 0 | 3614 | 0 | 3614 |
| 1995 | 2940 | 183 | 532 | 0 | 3655 | 0 | 3655 |
| 1996 | 2354 | 148 | 370 | 0 | 2872 | 0 | 2872 |
| 1997 | 2553 | 102 | 391 | 0 | 3046 | 0 | 3046 |
| 1998 | 2023 | 68 | 350 | 0 | 2441 | 0 | 2441 |
| 1999 | 2792 | 56 | 409 | 0 | 3257 | 0 | 3257 |
| 2000 | 2695 | 28 | 524 | 0 | 3247 | 0 | 3247 |
| 2001 | 2649 | 42 | 568 | 0 | 3259 | 0 | 3259 |
| 2002 | 2775 | 79 | 586 | 0 | 3440 | 0 | 3440 |
| 2003 | 2606 | 45 | 618 | 0 | 3269 | 0 | 3269 |
| 2004 | 2391 | 30 | 661 | 0 | 3082 | 0 | 3082 |
| 2005 | 2270 | 23 | 656 | 0 | 2949 | 0 | 2949 |
| 2006 | 3446 | 23 | 697 | 0 | 4166 | 0 | 4166 |


| UK Scotland |  |  |  |  |  | Other United Kingdom and Ireland | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | Nephrops trawl | other | creel | Below Minimum Size | Subtotal |  |  |
| 2007 | 3361 | 26 | 591 | 0 | 3978 | 0 | 3978 |
| 2008 | 3229 | 13 | 557 | 0 | 3799 | 0 | 3799 |
| 2009 | 2849 | 34 | 613 | 0 | 3496 | 0 | 3496 |
| 2010 | 1783 | 9 | 621 | 0 | 2413 | 0 | 2413 |
| 2011 | 2109 | 17 | 571 | 0 | 2697 | 0 | 2697 |
| 2012 | 2963 | 12 | 565 | 0 | 3540 | 2 | 3542 |
| 2013 | 2356 | 480 | 575 | 0 | 3411 | 2 | 3413 |
| 2014 | 2752 | 13 | 490 | 0 | 3255 | 2 | 3257 |
| 2015 | 2561 | 23 | 418 | 0 | 3002 | 0 | 3002 |
| 2016 | 3039 | 15 | 475 | 0.4 | 3529.4* | 0 | 3529.4* |
| 2017 | 2086 | 30 | 374 | 0 | 2489 | 1 | 2490 |
| 2018 | 1592 | 30 | 331 | 0 | 1950 | 3 | 1953 |
| 2019 | 1521 | 31 | 425 | 0 | 1975 | 2 | 1977 |
| 2020 | 900 | 17 | 414 | 0 | 1331 | 0 | 1331 |
| 2021 | 1547 | 53 | 472 | 1.1* | 2073.1 | 0 | 2073.1 |

*Below minimum size landings not rounded to show it was reported.

Table 12.2.2. Nephrops Scottish sampling levels all FUs in 6.a (including N. Irish for Clyde).

|  |  | 2019 |  | 2020 |  | 2021 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FU |  | N trips* | $N$ measured | N trips* | $N$ measured | $N$ trips* | $N$ measured |
| North Minch | Landings | 41 | 23952 | 25 | 8551 | 34 | 13368 |
|  | Discards | 35 | 3658 | 4 | 443 | 9 | 1439 |
| South Minch | Landings | 40 | 21378 | 18 | 8203 | 33 | 13770 |
|  | Discards | 25 | 1578 | 7 | 673 | 3 | 306 |
| Clyde | Landings | 22 | 19227 | 24 | 10037 | 31 | 14510 |
|  | N.Irish La |  |  |  |  |  |  |
|  | Discards | 33 | 4073 | - | - |  |  |

[^6]Table 12.2.3. Nephrops mean weight in the landings (FU11-13).

| Year | FU11 | FU12 | FU13 |
| :---: | :---: | :---: | :---: |
| 1990 | 21.39 | 19.99 | 24.27 |
| 1991 | 25.35 | 21.74 | 20.65 |
| 1992 | 21.66 | 24.10 | 25.16 |
| 1993 | 20.79 | 21.26 | 29.44 |
| 1994 | 23.45 | 24.96 | 25.28 |
| 1995 | 22.24 | 21.96 | 19.24 |
| 1996 | 26.68 | 23.10 | 21.68 |
| 1997 | 21.71 | 23.37 | 24.21 |
| 1998 | 23.65 | 22.18 | 17.98 |
| 1999* | 22.70 | 25.14 | 17.39 |
| 2000 | 24.19 | 27.30 | 19.96 |
| 2001 | 25.33 | 23.79 | 19.46 |
| 2002 | 25.93 | 26.83 | 16.35 |
| 2003 | 26.03 | 27.86 | 19.13 |
| 2004 | 25.16 | 27.37 | 18.80 |
| 2005 | 27.65 | 28.11 | 17.96 |
| 2006 | 24.52 | 26.24 | 19.27 |
| 2007 | 23.61 | 23.95 | 19.05 |
| 2008 | 23.90 | 23.91 | 16.59 |
| 2009 | 25.42 | 23.87 | 18.31 |
| 2010 | 29.39 | 25.86 | 21.21 |
| 2011 | 27.56 | 31.10 | 19.34 |
| 2012 | 23.43 | 29.17 | 21.83 |
| 2013 | 27.52 | 27.48 | 20.72 |
| 2014 | 27.96 | 29.91 | 20.79 |
| 2015 | 28.74 | 28.15 | 22.21 |
| 2016 | 25.76 | 24.76 | 17.70 |
| 2017 | 25.89 | 27.76 | 17.02 |
| 2018 | 27.39 | 27.27 | 16.14 |
| 2019 | 26.59 | 28.54 | 17.2 |
| 2020 | 31.06 | 36.58 | 18.96 |
| 2021 | 34.78 | 29.96 | 15.27 |
| Average** | 26.54 | $27.43$ | 16.91 |

* From 1999 onwards, mean weights are shown for trawl and creels combined.
** Average for FU11 and FU12 (1999-2021); FU13 (2018-2021).

Table 12.2.4. Nephrops, North Minch (FU11): Adjusted TV survey abundance, landings, discard rate (proportion by number) and estimated harvest rate.

| YEAR | LANDINGS IN NUMBERS (MILLIONS) | DISCARDS IN NUMBERS (MILLIONS) | REMOVALS IN NUMBERS (MILLIONS)** | ADJUSTED SURVEY VMS (MILLIONS)* | HARVEST RATE VMS | LANDINGS (TONNES) | DISCARDS (TONNES) | DISCARD <br> RATE | DEAD DISCARD RATE | MEAN WEIGHT IN LANDINGS (g) | MEAN WEIGHT IN DISCARDS (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 144 | 28 | 165 | 794 | 20.7 | 3257 | 273 | 16.4 | 12.8 | 22.7 | 9.69 |
| 2000 | 134 | 10 | 142 | 1166 | 12.1 | 3247 | 100 | 6.9 | 5.2 | 24.19 | 10.08 |
| 2001 | 129 | 17 | 141 | 1092 | 13 | 3259 | 160 | 11.7 | 9.1 | 25.33 | 9.32 |
| 2002 | 133 | 28 | 154 | 1337 | 11.5 | 3440 | 277 | 17.6 | 13.8 | 25.93 | 9.78 |
| 2003 | 126 | 30 | 148 | 1751 | 8.5 | 3269 | 299 | 19.2 | 15.2 | 26.03 | 10 |
| 2004 | 122 | 18 | 136 | 1751 | 7.8 | 3082 | 202 | 13 | 10.1 | 25.16 | 11.02 |
| 2005 | 107 | 50 | 144 | 1540 | 9.4 | 2949 | 507 | 32 | 26.1 | 27.65 | 10.09 |
| 2006 | 170 | 74 | 225 | 1762 | 12.8 | 4166 | 757 | 30.3 | 24.6 | 24.52 | 10.27 |
| 2007 | 168 | 12 | 177 | 1206 | 14.7 | 3978 | 214 | 6.5 | 5 | 23.61 | 18.1 |
| 2008 | 159 | 19 | 173 | 1047 | 16.5 | 3799 | 194 | 10.5 | 8.1 | 23.9 | 10.36 |
| 2009 | 138 | 35 | 164 | 1195 | 13.7 | 3496 | 327 | 20.3 | 16 | 25.42 | 9.34 |
| 2010 | 82 | 12 | 91 | 1293 | 7 | 2413 | 128 | 12.4 | 9.6 | 29.39 | 10.98 |
| 2011 | 96 | 16 | 108 | 1726 | 6.3 | 2697 | 154 | 14.2 | 11 | 27.56 | 9.66 |
| 2012 | 151 | 21 | 167 | 891 | 18.7 | 3542 | 213 | 12.6 | 9.3 | 23.43 | 10.33 |
| 2013 | 122 | 24 | 140 | 1403 | 10 | 3413 | 364 | 16.4 | 12.8 | 27.52 | 15.18 |
| 2014 | 115 | 8 | 121 | 1251 | 9.6 | 3257 | 77 | 6.3 | 4.8 | 27.96 | 9.99 |
| 2015 | 103 | 15 | 114 | 1445 | 7.9 | 3002 | 143 | 12.6 | 9.8 | 28.74 | 9.66 |
| 2016 | 136 | 22 | 152 | 1422 | 10.7 | 3529*** | 266 | 14 | 10.9 | 25.76 | 12.05 |
| 2017 | 95 | 5 | 99 | 1050 | 9.4 | 2491 | 65 | 5.3 | 4 | 25.89 | 12.51 |
| 2018 | 72 | 5 | 75 | 1188 | 6.4 | 1956 | 59 | 6.6 | 5.1 | 27.39 | 11.46 |
| 2019 | 74 | 4 | 78 | 1232 | 6.3 | 1979 | 51 | 5.5 | 4.2 | 26.59 | 11.92 |
| 2020 | 43 | 3 | 45 | 1439 | 3.1 | 1331 | 31 | 5.7 | 4.3 | 31.06 | 11.84 |
| 2021 | 61 | 4 | 64 | 1391 | 4.6 | 2073.1 | 65 | 6.2 | 4.7 | 34.78 | 16.02 |
| 2022 |  |  |  | 1346 |  |  |  |  |  |  |  |
| Average**** |  |  |  |  |  |  |  |  | 4.6 | 26.54 | 11.28 |

${ }^{*}$ harvest rates previous to 2006 are unreliable.
${ }^{* *}$ Removals numbers take the dead discard rate into account.
${ }^{* * *}$ Includes 0.4 tonnes of below minimum size landings.
**** Dead discard average: 2018-2021; Mean weight in landings and discards average: 1999-2021.

Table 12.2.5. Nephrops, North Minch (FU11): Results of the 1994-2022 TV surveys (values adjusted for bias).

| YEARS | NUMBER OF VALID STATIONS | MEAN DENSITY (BURROWS/ $\mathrm{M}^{2}$ ) | ABUNDANCE (SEDIMENT; MILLIONS) | 95\% CONFIDENCE INTERVAL (SEDIMENT; MILLIONS) | ABUNDANCE <br> (VMS; MILLIONS) | 95\% CONFI- <br> DENCE INTERVAL <br> (VMS; MILLIONS) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 41 | 0.29 | 500 | 74 | 820 | 122 |
| 1995 |  |  |  | No Survey |  |  |
| 1996 | 38 | 0.19 | 330 | 47 | 541 | 76 |
| 1997 |  |  |  | No Survey |  |  |
| 1998 | 38 | 0.31 | 547 | 77 | 898 | 127 |
| 1999 | 36 | 0.27 | 484 | 89 | 794 | 147 |
| 2000 | 39 | 0.40 | 711 | 82 | 1166 | 134 |
| 2001 | 56 | 0.38 | 666 | 81 | 1092 | 133 |
| 2002 | 37 | 0.46 | 815 | 91 | 1337 | 149 |
| 2003 | 41 | 0.60 | 1068 | 129 | 1751 | 211 |
| 2004 | 38 | 0.60 | 1068 | 107 | 1751 | 175 |
| 2005 | 41 | 0.53 | 939 | 100 | 1540 | 164 |
| 2006 | 30 | 0.61 | 1074 | 101 | 1762 | 165 |
| 2007 | 36 | 0.41 | 735 | 92 | 1206 | 150 |
| 2008 | 41 | 0.36 | 638 | 95 | 1047 | 157 |
| 2009 | 26 | 0.41 | 729 | 138 | 1195 | 227 |
| 2010 | 37 | 0.44 | - | - | 1293 | 231 |
| 2011 | 41 | 0.59 | - | - | 1726 | 226 |
| 2012 | 41 | 0.31 | - | - | 891 | 181 |
| 2013 | 41 | 0.48 | - | - | 1403 | 206 |
| 2014 | 44 | 0.43 | - | - | 1251 | 171 |
| 2015 | 41 | 0.50 | - | - | 1445 | 370 |
| 2016 | 39 | 0.49 | - | - | 1422 | 290 |
| 2017 | 42 | 0.36 | - | - | 1050 | 149 |
| 2018 | 44 | 0.40 | - | - | 1188 | 244 |
| 2019 | 47 | 0.42 | - | - | 1232 | 256 |
| 2020 | 33 | 0.49 | - | - | 1439 | 319 |
| 2021 | 50 | 0.48 | - | - | 1391 | 215 |
| 2022 | 36 | 0.46 |  |  | 1346 | 355 |

Table 14.2.6. Nephrops, North Minch (FU11): Results of the 2022 TV survey.

| STRATUM | AREA <br> $\left(\mathrm{km}^{2}\right)$ | NUMBER <br> OF STA- <br> TIONS | MEAN <br> BURROW <br> DENSITY <br> $\left(\mathrm{no}. / \mathrm{m}^{2}\right)$ | OB- <br> SERVED <br> VARI- <br> ANCE | ABUN- <br> DANCE <br> (MILLIONS) | STRATUM <br> VARI- <br> ANCE | PROPORTION <br> OF TOTAL <br> VARIANCE | SURVEY <br> PRECISION <br> LEVEL (CV) |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2022 TV survey |  |  |  |  |  |  |  |  |  |
| VMS | 2908 | 36 | 0.463 | 0.134 | 1345.9 | 31531 | 1 |  |  |
| Total | 2908 | 36 |  |  | 1345.9 | 31531 | 1 | 0.132 |  |



Figure 12.1. Nephrops Functional Units in 6.a. North Minch (FU11), South Minch (FU12), Clyde (FU13).


Figure 12.1.1. Nephrops in Division 6.a. Landings (tonnes) by functional unit (FU11, 12 \&13) and from rectangles outside the functional units (6.a outside FU).

## Landings - International



Effort - Scottish Nephrops trawlers



Figure 14.2.2. Nephrops, North Minch (FU11), Catch length-frequency distribution (dotted) and landings (solid) for Nephrops, 2000-2021. Mean size in catches and landings are represented by solid and dashed lines, respectively. Vertical dotted lines are minimum conservation reference size ( 20 mm ) and 35 mm .


Figure12.2.3 (a). Nephrops, North Minch (FU11), Proportion of landed weight by sex (top), by quarter (bottom) from Scottish trawlers.


Figure 12.2.3 (b). Nephrops, North Minch (FU11), quarterly numeric proportions by sex (2007-2021).


Figure 12.2.4. Nephrops, (FU11 North Minch, FU12 South Minch and FU13 Clyde), mean weight in the landings from 19902021 (from Scottish market sampling data).


Figure 12.2.5. Nephrops, (FU11 North Minch, FU12 South Minch, FU13 Clyde), mean weight in 2011-2021 by sample date, sex, métier and functional unit.


Figure 12.2.6. Nephrops, North Minch (FU11), TV survey station distribution and density (mean burrows/m²), 2016-2021. Bubbles in these figures are all scaled the same. Red crosses represent zero observations.


Figure 12.2.7. Nephrops, North Minch (FU11), time-series of revised TV survey abundance estimates (adjusted for bias; solid black line), with 95\% confidence intervals (dashed black lines), 1994-2022 (no survey in 1995 and 1997). The dashed red line is the rounded $\mathrm{B}_{\text {trigger }}$ value of $\mathbf{5 4 0}$ million individuals.


Figure 12.5.1. Nephrops, North Minch (FU11), harvest rate, 1994-2021 (no survey data in 1995 and 1997). The harvest rate is calculated by dead removals/TV abundance. The dashed and solid lines are the Fmsy proxy harvest rate $(10.8 \%)$ and the time-series of estimated harvest rates, respectively. Harvest rates prior to 2006 are considered unreliable.


Figure 12.6.1. The area of Nephrops habitat (estimated from VMS data) within the North Minch (FU11) relative to the areas of the Nature Conservation MPA (NCMPA) and Special Area of Conservation (SAC) showing areas within these where demersal trawling is banned (hatched) and where it is permitted for vessels below 500 kW (clear; depending on gear type, see SG, 2016). Geographic Coordinate System: OSGB 1936, Datum: OSGB 1936, Projected Coordinate System: British National Grid. Coastline by Wessel and Smith (2016), MPA sites subsetted from NCMPA (SNH, 2015) and SAC (SNH, 2016) layers, management areas by SG (2017b) and functional units generated from merged ICES rectangles (ICES, 2017). Map and modified layers created using ArcGIS (ESRI, 2014).

# 13 Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 12 (West of Scotland, South Minch) 

## Type of assessment in 2022

The assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follow the process defined by the benchmark WG (WKNEPH, 2009; WKNEPH, 2013). Full details are provided in the stock annex.

## ICES advice applicable to 2021

'ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, catches in 2021 that correspond to the F ranges in the MAP are between 4703 tonnes and 5916 tonnes, assuming recent discard rates. The entire range is considered precautionary when applying the ICES advice rule.

To ensure that the stock in Functional Unit 12 is exploited sustainably, management should be implemented at the functional unit level.'

## ICES advice applicable to 2022

'ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2018-2020, catches in 2022 should be no more than 3977 tonnes.

To ensure that the stock in Functional Unit (FU) 12 is exploited sustainably, management should be implemented at the FU level.

ICES notes the existence of a management plan, developed and adopted by one of the relevant management authorities for Subarea 6. ICES considers this plan to be precautionary when implemented at the FU level.'

### 13.1 General

## Stock description

The South Minch (FU12) is located midway down the west coast of Scotland (see. Section 12 FU11 North Minch, Figure 12.1). The area is characterised by numerous islands of varying size, with sea lochs occurring along the mainland coast. These topographical features create a diverse habitat with complex hydrography and a patchy distribution of various soft sediments. Further details are provided in the stock annex.

## Management applicable to 2021 and 2022

Management is at the ICES subarea level as described at the beginning of Section 12 FU11 North Minch.

## Ecosystem aspects

Details of the ecosystem aspects for this functional unit are provided in the stock annex where available.

## Fishery description

Information on developments in the fishery was provided by Marine Scotland compliance officers. In 2021 the fishery was described as remaining generally very poor. The COVID-19 affected the markets for product, with even small local markets ceasing operations due to closure of restaurants and other establishments for extended periods of time.
Two distinct fleets operate in the South Minch and the main ports are Oban and Mallaig. In Oban there are 56 local vessels ( $40<10 \mathrm{~m}$ vessels), while there was no information available from Mallaig for 2021. The local fleet in Oban changes quite frequently, e.g. some operators attempt to move to larger vessels but cannot find sufficient numbers to crew them.

Since 2009, vessels have been required to fit 120 mm square meshed panels, in accordance with the west coast emergency measures (Council Reg. (EU) 43/2009). Large SMPs ( 200 mm ) are also widely used and were made mandatory for all TR2 vessels with power $>112 \mathrm{~kW}$ fishing as part of the previous Scottish Conservation Credits scheme. Twin rig vessels tend to use a 200 mm square mesh panel with a 100 mm or larger mesh codend. These vessels do not catch bulk quantities and this leads to Nephrops of better average size and quality. A comment was noted in 2017 about the use of bungee cords to keep the meshes closed. This was investigated by Compliance officers but was deemed to be legal and was not reported as a problem in subsequent years.

There is very little fish bycatch landed due to the restrictions on cod, haddock and whiting (detailed in ICES, 2016a, ICES, 2016b and ICES, 2016c). Estimates of discard rates of haddock and whiting remain high (ICES, 2016d and ICES, 2017a). Haddock in areas 6 a are now covered by the landings obligation in area.

Further general information on the fishery can be found in the stock annex.

### 13.2 Data available

## InterCatch

Data for 2021 were uploaded to InterCatch prior to the 2022 WG meeting. Uploaded data were worked up in InterCatch to generate 2021 raised international length-frequency distributions. Allocation schemes for any unsampled fleets are described in the stock annex. Data exploration in InterCatch has previously shown that outputs of raised data were very close to those generated by the previous method applied internally with differences being $<0.1 \%$. As such, InterCatch length-frequency outputs have been used in the stock assessment since 2012.

The COVID-19 pandemic resulted in reduced sampling of commercial catches for FU12 in 2021, as was also the case in 2020. Discard sample data for FU12 was only available for quarter three of 2021, and so, InterCatch estimates of discard rates for quarters one, two, and four were based on samples collected in quarter three. Following download of data from InterCatch, alternate methods of 2021 discard estimation were also considered. It was agreed at WGCSE that estimates of discard rates and associated size distributions for 2021 would be based on an averaging of discard samples across all available quarters between 2017 and 2021. Minimum and maximum discard rates over the same period were also examined to gain an appreciation of the plausible range of discarding that might have occurred in 2021. Assessment estimates affected by changes
in discard rates are annotated hereafter to reflect this; i.e. " $x(y / z)$ ", where $x$ is the estimate based on the average discard rate from samples available between 2017 and 2021, $y$ is based on minimum discard rate, and $z$ on maximum discard rate over the same period.

## Commercial catch

Official catch statistics (landings) reported to ICES are shown in Table 12.1.1 (see. Section 12 FU11 North Minch). These relate to the whole of $6 . a$, of which the South Minch is a part. Landings for FU12 provided through national laboratories are presented in Table 13.2.1, broken down by country and by gear type. Landings from this fishery are predominantly reported from Scotland, with low levels reported from the rest of the UK and Ireland. Total reported Scottish landings in 2021 were 2577.3 tonnes (plus 77 tonnes from other UK vessels, and 42 tonnes from Ireland), consisting of 1939 tonnes ( $75 \%$ ) landed by Scottish Nephrops trawlers and 619 tonnes (24\%) landed by Scottish creel vessels. A small amount of below minimum size (BMS) landings was also reported ( 0.3 tonnes). The proportion of creel caught landings has generally increased somewhat over the past decade, from $19 \%$ in 2012 to $24 \%$ in 2021.

## Effort data

In 2015, WGCSE agreed that effort should be reported in kW days as this is likely to be more informative about changes in the actual fleet effort. Effort shows an overall decreasing trend since 2003 but there are peaks in 2008 and 2012, which can be attributed to visiting North Sea trawlers (Figure 16.2.1). The decline in effort observed in recent years halted in 2021, with a slight observed increase in effort following the extensive interruptions the fishery experienced in 2020 due to the COVID-19 pandemic. Note that the effort time-series range (2000-2021) does not match the more extensive range available for landings due to a lack of confidence in the reliability of older effort data in the Marine Scotland Science database. The effort is also slightly inconsistent with the landings data because effort is provided for TR2 vessels only, while the 'Nephrops trawl' landings also includes landings by large mesh trawlers targeting Nephrops.

## Sampling levels

Length compositions of landings and discards are obtained during market sampling and onboard observer sampling respectively. These sampling levels are shown in Table 12.2 .2 (see. Section 12 FU11 North Minch). Sampling effort in 2021 was lower than years preceding 2020, due to disruptions to both the fishing industry and government sampling programmes attributable to the COVID-19 pandemic. Length compositions for the creel fishery are available for landings only as the small numbers of discards from the fishery survive well. Therefore, these animals are not considered to be removed from the population, and $100 \%$ survival is assumed (ICES, 2013).

## Length compositions

Figure 13.2.2 shows a series of annual length-frequency distributions from 2002 onwards which appear fairly stable across the time-series. Catch and landings length compositions, and mean size are shown for each sex. The mean size of males and females decreased slightly in 2021.

## Sex ratio

The sex ratios observed in the South Minch in $2020 \& 2021$ showed some divergence from those observed in years with better sample coverage, although males still made the largest contribution
to the annual landings in most quarters. In the years prior to 2020, males were available to the fishery throughout the year while females were mainly caught in the summer when they emerge from their burrows after egg hatching has occured. This seasonal change could be observed in the quarterly sex ratios, with males dominating the catch in quarters one and four, and a more even sex ratio observed in quarters two and three. However, in 2020, all quarterly sex ratios were majority male (Figure 13.2.3) due to the decreased number of samples which were available for the year. In 2021, the quarterly sex ratios were more similar to typical years, although quarter four had the highest male ratio of the series. This metric is used as an indicator, whereby increasing proportions of females in the catch might signal an effect of acute overfishing. In the case of recent years, however, the unusual sex ratios are known to be due to poor sampling, and not a cause for concern to management.

## Mean weights

The mean weight in the landings (Figures 15.2 .4 and 15.2.5; see. Section 12 FU11 North Minch, Table 15.2.3) have fluctuated around a relatively high level since 2011. Seasonal variability (and occasional outliers) in mean weights are seen in the individual sample estimates. There appears to be a small increase in the mean weight of the males for the trawl caught Nephrops and also for both males and females caught by creels (Figure 15.2.5). The annual estimate of mean weight in the landings has an effect on the catch forecast. Over the time-series, there is a general increasing trend in mean weights in the landings, however, there was a substantial decrease in mean weight in 2021 following the highest recorded mean weight in 2020.

## Discarding

Discarding of undersized and unwanted Nephrops occurs in this fishery. Discard sampling has been conducted on the Scottish Nephrops trawler fleet since 1990. Discarding rates in this FU have varied considerably over the years, ranging from $3 \%$ to over $25 \%$. In 2021, the discarding rate was $5.6 \%(2.1 \% / 28.6 \%)$ based on mean discard rates observed between 2017 and 2021. This represents a decrease on $2020(7.8 \%)$. The low levels of discarding in recent years may be explained by poor fishing and a gradually decreasing fleet (Table 13.2.2).
Studies (Charuau et al., 1982; Sangster et al., 1997; Wileman et al., 1999) suggest that some Nephrops survive the discarding process, thus, an estimate of $25 \%$ survival is assumed for this FU in order to calculate removals (landings + dead discards) from the population. The discard survival rate for creel caught Nephrops has been shown to be high (ICES, 2013) and a value of $100 \%$ is assumed. The discard rate (adjusted to account for survival) which will be used in the forecast was estimated by taking a three-year average (2019-2021), amounting to $6.8 \%$ (3.7\%/10.7\%).

## Abundance indices from UWTV surveys

An underwater TV survey of the stock is conducted annually according to standards set out by the Manual for the Nephrops Underwater TV Surveys (Dobby et al., 2021). Surveys use a stratified random approach, and have been carried out for this stock since 1995. TV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows. The numbers of valid stations used in the final analysis in each year are shown in Table 16.2.3. On average, 36 stations have been considered valid each year, and raised to a stock area of $5072 \mathrm{~km}^{2}$ (derived from BGS sediment data). The 2022 UWTV survey of FU 12 was carried out in two stages, during June/July and August/September of 2022. The first part was the regularly scheduled annual survey, while the second was an ad hoc survey to collect data in areas which
were missed during the annual survey for logistic reasons. The first part of the survey collected random stations mostly located in the inshore areas to the east and south of the FU, and some in the central area. The second part of the survey collected stations in the northwest, some of which were randomly assigned and some which were moved from the southwest to maximise sample collection during the time which was at the survey team's disposal. A total of 41 valid TV stations were completed in 2022 and used in the survey analysis (Table 13.2.3; 24 stations in the first part of the survey and 17 in the second part), with one station excluded from the second part of the survey due to an excessively rocky seabed.

TV survey abundance estimates from 1995-2022 are shown in Table 13.2.3 and Figure 13.2.4. Since 2007, the stock has undergone cycles wherein abundance oscillates between high and low values over five to six year periods (Figure 13.2.4), with changes of up to 1199 million individuals between the lowest and highest points of a cycle (between 2012 and 2016). The 2022 abundance represents a $32 \%$ increase in relation to 2021.

Table 13.2.4 shows a more detailed summary of the results from the three most recent TV surveys conducted in FU12. The table includes estimates of abundance and variability in each of the strata adopted in the stratified random approach. Mean burrow density decreased in 2022, in comparison to the 2021 survey, in all strata apart from Sandy Mud. Densities are generally lower in the western parts of the area towards the Outer Hebrides and higher in the inshore areas to the south west of Skye (Figure 13.2.5). The CV for the 2022 TV survey (Table 13.2.4) is lower than the $20 \%$ precision level agreed by WGNEPS (2019; 12.6\%).

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009; ICES, 2013), WGNEPS (ICES, 2018a), WKNEPS (ICES, 2018b) and (Leocádio et al., 2018). A number of potential biases were highlighted including those due to edge effects, species burrow misidentification and burrow occupancy. The cumulative relative to absolute conversion factor estimated for FU12 was 1.32 meaning that the TV survey is likely to overestimate Nephrops abundance by $32 \%$.

### 13.3 Assessment

## Comparison with previous assessments

The assessment follows the same procedure as last year and is based on a combination of examining trends in fishery indicators and underwater TV abundance estimates. The process was defined by the benchmark WG and is described in the stock annex.

## State of the stock

The underwater TV survey is presented as the best available information on the South Minch (FU12) Nephrops stock. The details of the 2022 survey are shown in Table 16.2.4, and compared with the 2020 and 2021 outcomes. At present, it is not possible to extract any length or age structure information from the survey and therefore it only provides information on abundance over the area of the survey.

TV survey estimated stock abundance in 2022 was 1677 million individuals, above the MSY Btrigger value of 1016 million, or the rounded value of 1020 million used for the provision of advice.

The calculated harvest rate in 2021 (dead removals/TV abundance $=7.5 \% ; \underline{7.2 \% / 9.2 \%}$ ) was below the Fmsy proxy for this stock (the value associated with high long-term yield and low risk depletion) of $11.7 \%$.

### 13.4 Catch scenarios table

Landings predictions and catch options at various harvest rates (based on principles established at WKNEPH (ICES, 2009), are made on the basis of the 2022 UWTV survey conducted in June. These were presented at WGCSE NEPH in September 2022 for the provision of advice.

Catch scenarios table inputs and historical estimates of mean weight in landings and harvest rates are presented in Table 16.2.2 and summarised below. The calculation of catch scenarios for the South Minch follows the procedure outlined in the stock annex.

Given the variability in mean weights it was considered more appropriate to use a full timeseries average, from 1999 (first year with creel and trawl length distributions combined) until 2021.

The table below shows the agreed inputs to the catch scenario table.

| Input | Data | 2022 assessment |
| :--- | :---: | :---: |
| Survey abundance (millions) | UWTV 2022 | 1677 |
| Mean weight in projected landings (g) | $1999-2021$ | 27.43 |
| Mean weight in projected discards (g) | $1999-2021$ | 10.18 |
| Projected discard rate | $2019-2021$ | $6.8 \%^{*}$ |
| Discard survival rate | Proportion by number (assumed) | $25 \%^{* *}$ |

* Based on mean discard rate (2017-2021) allocated to 2021; estimates of $10.7 \%$ and $3.7 \%$ were derived based on the maximum and minimum observed discard rates, respectively, for the same period
** Discard survival in the creel fishery is assumed to be $100 \%$, as outlined in the stock annex.


### 13.5 Reference points

New reference points were derived for this stock at WKMSYRef4 (ICES, 2016e). These are updated on the basis of an average of estimated Fmsy proxy harvest rates over a period of years which corresponds more closely to the methodology for finfish. In cases where there is a clear trend in the values, a five-year average was chosen. Similarly, the five-year average of the F at $95 \%$ of the YPR obtained at the Fmsy proxy reference point was proposed as the Fmsy lower bound and the five-year average of the F above Fmax that leads to YPR of $95 \%$ of the maximum as the upper bound. Using an average value also has the advantage of reducing the effect of any unusually high or low estimates of the Fmsy proxy, which occasionally appear. For this stock, the Fmsy proxy has been revised from $12.3 \%$ to $11.7 \%$.
For Nephrops stocks, MSY Btrigger has been defined as the lowest stock size from which the abundance has increased and is calculated as 1016 million individuals (in 2010). This value was rounded to 1020 million, in the advice from WKMSYRef4 on MSY Btrigger. Full details are contained in the stock annex.

These should remain under review by WGCSE and may be revised should improved data become available.

Table 13.2.2 and Figure 13.5.1 show the harvest rates for FU12. The harvest rate has fluctuated over the time-series and has been below the Fmsy proxy since 2013. The increase in 2016, compared to the 2013-2015 harvest rates, was due to relatively increased landings compared to abundance. The harvest rate more than halved in 2018 compared to 2017, and has continued to decrease through 2020 to a new historical low, increasing again in 2021 (7.5\%; 7.2\%/9.2\%).

It is likely that prior to 2006, the harvest rates are underestimates due to under-reported landings.

### 13.6 Management strategies

Scotland has established a network of regional Inshore Fisheries Groups (rIFGs), non-statutory bodies that aim to improve the management of Scotland's inshore fisheries out to six nautical miles, and to give commercial inshore fishermen a strong voice in wider marine management developments. The rIFGs will contribute to regional policies and initiatives relating to management and conservation of inshore fisheries, including impacts on the marine environment and the maintenance of sustainable fishing communities and measures designed to better conserve and sustainably exploit stocks of shellfish and sea fish (including salmon) in their local waters. Although no IFG proposals specific to the management of Nephrops fisheries have yet been adopted, some of the IFG management plans for the Scottish West Coast include spatial management of Nephrops fisheries and the introduction of creel limits.

On the 8th of February 2016, phase 1 of the fisheries management measures for inshore MPAs in Scottish waters came into force (SG, 2016). These measures relate to both Nature Conservation MPAs (NCMPAs; Marine (Scotland) Act and the UK Marine and Coastal Access Act) and Special Areas of Conservation (SACs; EC Habitats Directives - Council Directive 92/43/EEC) both of which have the aim of conserving biological diversity in Scottish waters and contribute to Scotland's MPA network (SG, 2017a). Although not specific to the management of the Nephrops fishery, they will influence spatial patterns of fishing for Nephrops where controls on the two main gear types, demersal trawls and creels, are implemented on Nephrops habitat. There are seven protected areas within the South Minch functional unit with fisheries management measures. MPAs on the main areas of Nephrops habitat include the Loch Sunart to the Sound of Jura NCMPA where demersal trawling is banned in some areas, i.e. zoned, and seasonal closures implemented in others, Loch Sunart NCMPA/SAC, where demersal trawling is banned and creeling is zoned, the East of Mingulay SAC, demersal trawling banned and creeling zoned, and the Trenish Isles SAC, demersal trawling banned. Another area is the Loch Duich, Long and Alsh NCMPA/SAC, covering some patches of muddy sediment, where demersal trawling is banned or temporally closed in other areas that extend beyond the MPA onto muddy sediment. Other areas include the Loch Creran SAC/NCMPA, demersal trawling banned and creeling zoned, and the Firth of Lorn SAC, which has the same management as the Loch Sunart to the Sound of Jura NCMPA. For the Firth of Lorn and Loch Creran, management was in place prior to 2016 (SG, 2016). An additional NCMPA, at Loch Carron, was designated using emergence powers in 2017 (SG, 2017b). The areas of the SACs and NCMPAs relative to the estimated Nephrops habitat within the South Minch functional unit are displayed in Figure 13.6.1.

### 13.7 Quality of assessment and forecast

The length and sex composition of the landings data is considered to be generally adequately sampled, sampling levels have remained relatively consistent over the past two years (see Section 16.2), with the exception of quarter 2 of 2020 where sampling efforts were disrupted by the COVID-19 pandemic. Discard sampling has been conducted for Scottish Nephrops trawlers in this fishery since 1990, and is considered to represent the trawl fishery adequately. The reduced sampling effort in 2021 meant that discard sample data were only available for Quarter three, and it was agreed at WGCSE that estimates of discard rates and size distributions for 2021 would be adequately approximated for the purpose of forecasting by averaging of discard samples across all available Quarters between 2017 and 2021. The landings length compositions from 1999 onwards are derived from both creel and trawl samples. The creel fishery, which accounts for an increasing proportion of the landings in recent years ( $\sim 24-29 \%$ in the past three years) and
increasingly operates over similar areas to trawling, exhibits a length composition composed of larger animals.

There are concerns over the accuracy of historical landings and effort data prior to 2006 when Buyers and Sellers legislation was introduced and the reliability began to improve. Because of this, the final assessment adopted is independent of official statistics. Harvest rates since 2006 are also considered more reliable due to more accurate landings data reported under new legislation. Incorporation of creel length compositions has also improved estimates of harvest rates.
Underwater TV surveys have been conducted for this stock every year since 1995. The number of valid stations in the survey has remained relatively stable throughout the time period. The survey is targeted at known areas of mud, sandy mud and muddy sand within the South Minch. The variance of density estimates in the South Minch is relatively high, particularly in the sandy mud strata, resulting in large confidence intervals and a greater uncertainty on the abundance estimates than in other FUs. This makes it difficult to determine which population changes are significant. Although the CV's have been smaller in recent years.

There is a need to explore options to implement further stratification for the South Minch survey area. In the provision of catch options based on the absolute survey estimates, additional uncertainties related to mean weight in the landings and the discard rates also arise. A three-year average (i.e. 2019-2021 for the 2022 assessment) of discard rates (adjusted to account for some survival of discarded animals) has been used in the calculation of catch options.

The cumulative relative to absolute conversion factor estimates for FU12 are largely based on expert opinion. The precision of these bias corrections cannot yet be characterised. The landings derived in the forecast (catch options table) are sensitive to the input dead discard rate and mean weights in landings, and this introduces uncertainties in the catch forecasts. Precision estimates are needed for these forecast inputs.

The overall area of the ground is estimated from the available BGS contoured sediment data and at present is considered to be a minimum estimate. Work is underway to improve the area estimation. VMS data linked to landings (from queries of the Scottish FIN database) suggest no major differences between areas fished and the mud sediment maps. Two other factors however, are likely to increase the estimate of ground area available for Nephrops and Nephrops directed fishing. Firstly, the inclusion of vessels smaller than 15 m would likely increase the fished area in some of the inshore locations and secondly, it is known that most of the sea lochs have areas of mud substrate and are typically fished by creel boats. In recent years, limited TV surveys have taken place in some of the sea lochs and attempts are being made to utilise these data to improve estimates of mud area and Nephrops abundance in the South Minch.

### 13.8 Recommendation for next benchmark

This stock was last benchmarked in 2009. WGCSE will keep the stock under close review and recommend future benchmark as required.

At WGCSE 2022 it was agreed that a benchmark/interbenchmark should be carried out on FU12 Nephrops, addressing the potential for provision of abundance estimates with reduced uncertainty using alternate estimation methods (REF WGCSE 2022 report).

### 13.9 Management considerations

ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES division level. Management at the functional unit level could provide controls to ensure effort and catch were in line with resources available.

Creel fishing takes place in this area but overall effort in terms of creel numbers is not known and measures to control numbers are not in place. There is a need to ensure that the combined effort from all forms of fishing is taken into account when managing this stock.

There is a bycatch of other species in the area of the South Minch and estimated discards of whiting and haddock by the TR2 fleet are high in area 6.a generally. It is important that efforts continue to ensure that unwanted bycatch is kept to a minimum in this fishery.

### 13.10 References

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Table 13.2.1. Nephrops, South Minch (FU12), ICES estimates of landings of Nephrops, 1981-2021.

| UK SCOTLAND |  |  |  |  |  | OTHER UK | IRELAND | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | NEPHROPS TRAWL | OTHER | CREEL | BELOW MINIMUM SIZE | SUB TOTAL |  |  |  |
| 1981 | 2966 | 254 | 432 | 0 | 3652 | 0 | 0 | 3652 |
| 1982 | 2925 | 206 | 421 | 0 | 3552 | 0 | 0 | 3552 |
| 1983 | 2595 | 362 | 456 | 0 | 3413 | 0 | 0 | 3413 |
| 1984 | 3229 | 477 | 594 | 0 | 4300 | 0 | 0 | 4300 |
| 1985 | 3096 | 424 | 488 | 0 | 4008 | 0 | 0 | 4008 |
| 1986 | 2694 | 288 | 502 | 0 | 3484 | 0 | 0 | 3484 |
| 1987 | 2928 | 418 | 546 | 0 | 3892 | 0 | 0 | 3892 |
| 1988 | 3544 | 364 | 555 | 0 | 4463 | 10 | 0 | 4473 |
| 1989 | 3846 | 338 | 561 | 0 | 4745 | 0 | 0 | 4745 |
| 1990 | 3732 | 263 | 435 | 0 | 4430 | 0 | 0 | 4430 |
| 1991 | 3596 | 342 | 503 | 0 | 4441 | 1 | 0 | 4442 |
| 1992 | 3478 | 209 | 549 | 0 | 4236 | 1 | 0 | 4237 |
| 1993 | 3609 | 194 | 650 | 0 | 4453 | 5 | 0 | 4458 |
| 1994 | 3742 | 264 | 405 | 0 | 4411 | 3 | 0 | 4414 |
| 1995 | 3443 | 717 | 508 | 0 | 4668 | 14 | 0 | 4682 |
| 1996 | 3108 | 417 | 469 | 0 | 3994 | 1 | 0 | 3995 |
| 1997 | 3518 | 329 | 493 | 0 | 4340 | 3 | 1 | 4344 |
| 1998 | 2851 | 340 | 538 | 0 | 3729 | 0 | 1 | 3730 |
| 1999 | 3165 | 359 | 514 | 0 | 4038 | 0 | 14 | 4052 |
| 2000 | 2940 | 311 | 700 | 0 | 3951 | 0 | 2 | 3953 |
| 2001 | 2823 | 391 | 768 | 0 | 3982 | 0 | 9 | 3991 |
| 2002 | 2234 | 314 | 743 | 0 | 3291 | 0 | 14 | 3305 |
| 2003 | 2812 | 203 | 858 | 0 | 3873 | 0 | 6 | 3879 |
| 2004 | 2864 | 105 | 879 | 0 | 3848 | 0 | 21 | 3869 |
| 2005 | 2812 | 46 | 955 | 0 | 3813 | 1 | 34 | 3848 |
| 2006 | 3570 | 97 | 922 | 0 | 4589 | 9 | 35 | 4633 |
| 2007 | 4437 | 21 | 959 | 0 | 5417 | 19 | 35 | 5471 |


| UK SCOTLAND |  |  |  |  |  | OTHER UK | IRELAND | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | NEPHROPS TRAWL | OTHER | CREEL | BELOW MINIMUM SIZE | SUB TOTAL |  |  |  |
| 2008 | 4433 | 12 | 896 | 0 | 5341 | 2 | 13 | 5356 |
| 2009 | 3346 | 24 | 900 | 0 | 4270 | 4 | 11 | 4285 |
| 2010 | 2836 | 19 | 969 | 0 | 3824 | 16 | 6 | 3846 |
| 2011 | 2876 | 11 | 783 | 0 | 3670 | 23 | 9 | 3702 |
| 2012 | 3159 | 32 | 773 | 0 | 3964 | 19 | 6 | 3989 |
| 2013 | 2490 | 543 | 729 | 0 | 3762 | 13 | 1 | 3776 |
| 2014 | 2490 | 3 | 637 | 0 | 3130 | 32 | 17 | 3179 |
| 2015 | 2662 | 18 | 665 | 0 | 3345 | 22 | 33 | 3400 |
| 2016 | 3450 | 22 | 838 | 0 | 4310 | 33 | 59 | 4402 |
| 2017 | 2833 | 60 | 775 | 0 | 3668 | 23 | 66 | 3757 |
| 2018 | 1693 | 86 | 682 | 0 | 2461 | 45 | 34 | 2540 |
| 2019 | 1493 | 39 | 621 | 0 | 2153 | 29 | 38 | 2220 |
| 2020 | 1320 | 25 | 554 | 0 | 1899 | 8 | 69 | 1976 |
| 2021 | 1939 | 19 | 619 | 0.3 | 2577.3 | 77 | 42 | 2696.3 |

Table 13.2.2. Nephrops, South Minch (FU12): Adjusted TV survey abundance, landings, discard rate proportion by number) and estimated harvest rate.

| YEAR | LANDINGS NUMBER (MILLIONS) |  | REMOVALS NUMBER (MILLIONS)** | ADJUSTED SURVEY (MILLIONS) | HARVEST RATE* | LANDINGS (TONNES) | DISCARDS (TONNES) | DIS- <br> CARD <br> RATE <br> (\%) | DEAD DISCARD RATE (\%) | MEAN WEIGHT IN LANDINGS (g) | MEAN WEIGHT IN DISCARDS <br> (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 161 | 29 | 183 | 1086 | 16.9 | 4052 | 206 | 15.4 | 12 | 25.14 | 7 |
| 2000 | 145 | 33 | 170 | 1854 | 9.2 | 3953 | 284 | 18.7 | 14.7 | 27.3 | 8.5 |
| 2001 | 168 | 65 | 216 | 2037 | 10.6 | 3991 | 591 | 27.9 | 22.5 | 23.79 | 9.11 |
| 2002 | 123 | 26 | 143 | 1899 | 7.5 | 3305 | 247 | 17.6 | 13.8 | 26.83 | 9.37 |
| 2003 | 139 | 38 | 168 | 2157 | 7.8 | 3879 | 381 | 21.3 | 16.9 | 27.86 | 10.1 |
| 2004 | 141 | 44 | 175 | 2558 | 6.8 | 3869 | 454 | 23.8 | 19 | 27.37 | 10.26 |
| 2005 | 137 | 49 | 174 | 2208 | 7.9 | 3848 | 452 | 26.5 | 21.2 | 28.11 | 9.17 |
| 2006 | 177 | 30 | 199 | 1845 | 10.8 | 4633 | 324 | 14.3 | 11.1 | 26.24 | 10.97 |
| 2007 | 228 | 66 | 278 | 1016 | 27.3 | 5471 | 903 | 22.4 | 17.8 | 23.95 | 13.73 |
| 2008 | 224 | 74 | 279 | 1608 | 17.4 | 5356 | 605 | 24.7 | 19.8 | 23.91 | 8.23 |
| 2009 | 179 | 26 | 199 | 1542 | 12.9 | 4285 | 216 | 12.5 | 9.6 | 23.87 | 8.44 |
| 2010 | 149 | 12 | 158 | 2076 | 7.6 | 3846 | 133 | 7.7 | 5.9 | 25.86 | 10.76 |
| 2011 | 118 | 11 | 126 | 1945 | 6.5 | 3702 | 92 | 8.2 | 6.3 | 31.1 | 8.78 |
| 2012 | 133 | 16 | 145 | 919 | 15.8 | 3989 | 145 | 10.8 | 8.3 | 29.17 | 9.05 |
| 2013 | 136 | 4 | 140 | 1718 | 8.1 | 3776 | 50 | 3.1 | 2.4 | 27.48 | 11.31 |
| 2014 | 105 | 19 | 120 | 2073 | 5.8 | 3179 | 233 | 15.6 | 12.1 | 29.91 | 12.04 |
| 2015 | 120 | 10 | 128 | 1998 | 6.4 | 3400 | 121 | 7.7 | 5.9 | 28.15 | 12.04 |
| 2016 | 177 | 31 | 201 | 2118 | 9.5 | 4402 | 365 | 14.9 | 11.6 | 24.76 | 11.74 |
| 2017 | 131 | 13 | 140 | 1384 | 10.1 | 3757 | 108 | 9.4 | 7 | 27.76 | 8.29 |
| 2018 | 91 | 4 | 94 | 1946 | 4.8 | 2540 | 54 | 4.5 | 3.4 | 27.27 | 12.74 |
| 2019 | 79 | 4 | 83 | 2362 | 3.5 | 2220 | 46 | 4.9 | 3.7 | 28.54 | 11.22 |
| 2020 | 54 | 5 | 57 | 1927 | 3 | 1976 | 46 | 7.8 | 6 | 36.58 | 9.91 |
| 2021 | 90 | 7 | 95 | 1272 | 7.5 | 2696.3 | 84 | 7.6 | 5.8 | 29.96 | 11.35 |
| 2022 | - | - | - | 1677 | - | - | - | - | - | - | - |
| Average*** |  |  |  |  |  |  |  |  | 5.2 | 27.43 | 10.18 |

* Harvest rates previous to 2006 are unreliable.
** Removals numbers take the dead discard rate into account.
*** Dead discard average: 2019-2021; Mean weight in landings and discards average: 1999-2021.

Table 13.2.3. Nephrops, South Minch (FU12): Results of the 1995-2022 TV surveys (adjusted for bias).

| YEAR | NUMBER OF VALID STATIONS | MEAN DENSITY (BURROWS/m²) | ABUNDANCE (MILLIONS) | 95\% CONFIDENCE INTERVAL (MILLIONS) |
| :---: | :---: | :---: | :---: | :---: |
| 1995 | 33 | 0.227 | 1152 | 251 |
| 1996 | 21 | 0.288 | 1473 | 530 |
| 1997 | 36 | 0.212 | 1086 | 185 |
| 1998 | 38 | 0.288 | 1452 | 232 |
| 1999 | 37 | 0.212 | 1086 | 260 |
| 2000 | 41 | 0.364 | 1854 | 348 |
| 2001 | 47 | 0.402 | 2037 | 459 |
| 2002 | 31 | 0.371 | 1899 | 567 |
| 2003 | 25 | 0.424 | 2157 | 756 |
| 2004 | 38 | 0.508 | 2558 | 473 |
| 2005 | 33 | 0.432 | 2208 | 740 |
| 2006 | 36 | 0.364 | 1845 | 598 |
| 2007 | 39 | 0.197 | 1016 | 155 |
| 2008 | 33 | 0.318 | 1608 | 415 |
| 2009 | 25 | 0.303 | 1542 | 634 |
| 2010 | 34 | 0.409 | 2076 | 665 |
| 2011 | 36 | 0.383 | 1945 | 778 |
| 2012 | 38 | 0.182 | 919 | 185 |
| 2013 | 38 | 0.339 | 1718 | 365 |
| 2014 | 36 | $0.409$ | 2073 | 530 |
| 2015 | 35 | 0.394 | 1998 | 514 |
| 2016 | 37 | $0.417$ | $2118$ | $440$ |
| 2017 | 41 | 0.273 | 1384 | 282 |
| 2018 | 39 | 0.383 | 1946 | 371 |
| 2019 | 40 | 0.466 | 2362 | 578 |
| 2020 | 40 | 0.38 | 1927 | 517 |
| 2021 | 41 | 0.251 | 1272 | 339 |
| 2022 | 41 | 0.33 | 1677 | 471 |

Table 13.2.4. Nephrops South Minch (FU12). Results by stratum of the 2020-2022 TV surveys. Note that stratification was based on a series of sediment strata (M - Mud, SM - Sandy mud, MS - Muddy sand).

| STRATUM | AREA <br> ( $\mathrm{km}^{2}$ ) | NUMBER OF STATIONS | MEAN BURROW DENSITY (no./m²) | OBSERVED <br> VARIANCE | ABUN- <br> DANCE <br> (MIL- <br> LIONS) | STRATUM VARIANCE | PROPOR- <br> TION OF TOTAL VARIANCE | SURVEY <br> PRECISION <br> LEVEL (CV) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2020 \text { TV }$ <br> Suvey |  |  |  |  |  |  |  |  |
| M | 303 | 2 | 0.193 | 0.008 | 58.6 | 381 | 0.006 |  |
| SM | 2741 | 20 | 0.474 | 0.148 | 1299.7 | 55679 | 0.834 |  |
| MS | 2028 | 18 | 0.281 | 0.047 | 569.2 | 10685 | 0.16 |  |
| Total | 5071 | 40 |  |  | 1927.4 | 66745 |  | 0.131 |
| $2021 \text { TV }$ <br> Suvey |  |  |  |  |  |  |  |  |
| M | 303 | 3 | 0.402 | 0.084 | 121.7 | 2564 | 0.089 |  |
| SM | 2741 | 17 | 0.261 | 0.046 | 716.9 | 20574 | 0.716 |  |
| MS | 2028 | 21 | 0.214 | 0.029 | 433 | 5580 | 0.194 |  |
| Total | 5071 | 41 |  |  | 1271.6 | 28719 | 0.999 | 0.126 |
| $2022 \text { TV }$ <br> Suvey |  |  |  |  |  |  |  |  |
| M | 303 | 4 | 0.317 | 0.057 | 95.8 | 1324 | 0.024 |  |
| SM | 2741 | 16 | 0.448 | 0.105 | 1228.9 | 49419 | 0.89 |  |
| MS | 2028 | 21 | 0.173 | 0.024 | 351.8 | 4766 | 0.086 |  |
| Total | 5071 | 41 |  |  | 1676.5 | 55508 | 1 | 0.129 |

## Landings - International




Figure 13.2.1. Nephrops, South Minch (FU12). Long-term landings and effort.


Figure 13.2.2. Nephrops. South Minch (FU12). Catch length-frequency distribution and mean size in catches (dotted) and landings (solid) for Nephrops in the North Minch, 2002-2021. Vertical dotted lines are minimum conservation reference size ( $\mathbf{2 0} \mathrm{mm}$ ) and 35 mm .


Figure 13.2.3. (a) Nephrops, South Minch (FU12). Proportion of landings by sex and quarter from Scottish trawlers.


Figure 13.2.3 (b) Nephrops, South Minch (FU12), Proportion of males by quarter (2007-2021).


Figure 16.2.4. Nephrops, South Minch (FU12), Time-series of TV survey abundance estimate (adjusted for bias, solid black line), with $95 \%$ confidence intervals (dashed black lines), 1995-2022. The dashed red line is the rounded $\mathrm{B}_{\text {trigger }}$ value of 1020 million individuals.


Figure 16.2.5. Nephrops, South Minch (FU12), TV survey station distribution and relative density (burrows/m²), 20112022. Shaded green and brown areas represent areas of suitable sediment for Nephrops. Bubbles in this figure are all scaled the same. Red crosses represent zero observations.


Figure 16.5.1. Nephrops, South Minch (FU12), harvest rate, 1995-2021. The dashed and solid lines are the Fmsy proxy harvest rate (11.7\%) and the time-series of estimated harvest rates, respectively. Harvest rates prior to 2006 are considered unreliable.


Figure 16.6.1. The area of Nephrops habitat (Mud, Muddy Sand and Sandy Mud) within the South Minch (FU12) relative to the areas of the Nature Conservation MPAs (NCMPAs) and Special Area of Conservations (SACs) with fisheries management measures. Areas where demersal trawling is prohibited, restricted (i.e. vessel size restrictions or seasonal closures) and where creeling is prohibited are displayed. For more detailed information see SG (2016). Geographic Coordinate System: OSGB 1936, Datum: OSGB 1936, Projected Coordinate System: British National Grid. Coastline by Wessel and Smith (2016), MPA sites subsetted from NCMPA (SNH, 2015) and SAC (SNH, 2016) layers, management areas from SG (2017c) and functional units generated from merged ICES rectangles (ICES, 2017b). Map and modified layers created using ArcGIS (ESRI, 2014).

# 14 Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 13 (West of Scotland, the Firth of Clyde and Sound of Jura) 

## Type of assessment in 2022

The assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the benchmark WG (WKNEPH, 2009; WKNEPH, 2013). Full details are provided in the stock annex.

## ICES advice applicable to 2021

'ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, catches in 2021 that correspond to the F ranges in the MAP are between 3638 tonnes and 5425 tonnes (3142-4791 tonnes for the Firth of Clyde and 496-634 tonnes for the Sound of Jura), assuming recent discard rates. The entire range is considered precautionary when applying the ICES advice rule.

To ensure that Nephrops stocks are exploited sustainably, management of Nephrops should be implemented at the functional unit level. In this particular functional unit (FU), additional measures should be implemented to ensure that landings taken in each subarea (the Firth of Clyde and the Sound of Jura) are in line with the advice.'

## ICES advice applicable to 2022

'ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2018-2020, catches in 2022 should be no more than 4235 tonnes ( 3607 tonnes for the Firth of Clyde and 628 tonnes for the Sound of Jura).

To ensure that the stock in Functional Unit (FU) 13 is exploited sustainably, management should be implemented at the FU level. In this particular FU, additional measures should be implemented to ensure that landings taken in each subarea (the Firth of Clyde and the Sound of Jura) are in line with the advice.

ICES notes the existence of a management plan, developed and adopted by one of the relevant management authorities for Subarea 6. ICES considers this plan to be precautionary when implemented at the FU level.'

### 14.1 General

## Stock description

The Clyde functional unit (FU13) is located in the southern waters off the west coast of Scotland (see. Section 12 FU11 North Minch, Figure 12.1). It is comprised of two distinct patches in the Firth of Clyde and the Sound of Jura, to the east and west of the Mull of Kintyre respectively. The hydrography of the two subareas differs, with the Sound of Jura characterised by stronger tidal currents and the Firth of Clyde exhibiting features of a lower energy environment with a shallow entrance sill. Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Within the two distinct patches, these substrates are distributed according to prevailing hydrographic and bathymetric conditions. The available
area of suitable sediment is smaller in the Sound of Jura, occupying only the deepest parts of the Sound, while in the Firth of Clyde these sediments predominate. Further details are provided in the stock annex.

## Management applicable to 2021 and 2022

Management is at the ICES subarea level as described at the beginning of Section 12 FU11 North Minch.

## Ecosystem aspects

Details of the ecosystem aspects for this functional unit are provided in the stock annex where available.

## Fishery description

Information on developments in the fishery was provided by Marine Scotland fishery compliance officers.

There are 26 Nephrops vessels including 20 trawlers and six creelers fishing out of Ayr. Six boats left the fleet in 2021 due to crewing difficulties. All trawlers use 80 mm single or twin rigs with square mesh panels (SMP) of at least 160 mm , in accordance with The Sea Fishing (Licences and Notices) (Scotland) Regulations 2011. Nephrops trawling vessels with power $>200 \mathrm{~kW}$, or $>12 \mathrm{~m}$, are required to use a 300 mm SMP.

The activity of Northern Irish vessels was not perceived to be high since 2017, when compared to previous years. Many vessels have moved to other areas where there was better fishing, some travelling as far away as Eyemouth, and vessels fishing in FU13 did not land locally instead going back to their home port because of better fuel and market prices.

Mobile gear is banned in the Inshore Clyde from Friday night to Sunday night as are vessels greater than 21 m in length. Most creel boats operating in the Clyde have two crew members and operate around 1000 creels. Creeling activity now takes place quite widely in the northern parts of the Firth operating on some of the same grounds but often taking place during the weekend trawling ban.
In terms of the influence of Marine Protected Area (MPA) management measures on the fishery, the South Arran Nature Conservation MPA (NCMPA) removed a large sea area for Nephrops trawlers to operate over. Trawlers which would have operated in this area were displaced to areas where they would not have targeted previously, or where they would have only operated in poor weather conditions. This allowed creelers to move into the areas were trawling was banned. There have been recent reports of increases in creel numbers in this area and this has resulted in gear conflict within the creel sector. The small area of the Upper Loch Fyne NCMPA closed to trawlers was reported to have had little impact.

Further general information on the fishery can be found in the stock annex.

### 14.2 Data available

## InterCatch

Commercial data for 2021 were uploaded to InterCatch prior to the 2022 WG meeting. Uploaded data were worked up in InterCatch to generate 2021 raised international length-frequency
distributions. Allocation schemes for any unsampled fleets are described in the stock annex. Data exploration in InterCatch has previously shown that outputs of raised data were very close to those generated by the previous method applied internally with differences being $<0.1 \%$. As such, InterCatch length-frequency outputs have been used in the stock assessment since 2012.

The COVID-19 pandemic resulted in a reduced sampling effort of commercial catches for FU13 in 2021, as was also the case in 2020. There were no representative discards samples collected for FU13 in 2021, thus alternate methods of estimating discards were explored. It was agreed at WGCSE that estimates of discard rates and size distributions for 2021 would be based on an averaging of discard samples across all quarters from 2017-2019, the same procedure that was carried out for 2020 discards. Minimum and maximum discard rates over the same period were also examined to gain an appreciation of the plausible range of discarding that might have occurred in 2021. Assessment estimates affected by changes in discard rates are annotated hereafter to reflect this; i.e. " $x(y / z)$ ", where $x$ is the estimate based on the average discard rate 2017-2019, $y$ is based on minimum discard rate, and $z$ on maximum discard rate over the same period.

## Commercial catch

Official catch statistics (landings) reported to ICES are shown in Table 12.1.1 (see. Section 12 FU11 North Minch). These relate to the whole of area 6 .a of which the FU13 is a part. Landings statistics for FU13 provided through national laboratories are presented in Table 14.1.1, broken down by country and by gear type. Landings from this fishery are predominantly reported from Scotland, although Northern Ireland contributed 904 tonnes in 2021. Total reported Scottish landings in 2021 were 4088 tonnes (plus 907 tonnes from other UK vessels i.e. Northern Ireland \& England), consisting of 3805 tonnes landed by trawlers ( $93 \%$ ) and 233 tonnes ( $5.7 \%$ ) landed by Scottish creel vessels. Scottish creel landings have generally increased in the most recent years, from approximately 3\% in 2012 to just below 6\% of total landings in 2021.

Statistical rectangle 40E4 covers parts of both the Firth of Clyde and the Sound of Jura. Table 14.2.1 shows the split in landings between the two subareas comprising FU13. Historically the allocation of landings to the two components of FU13 was carried out by the fishery office and required them to have detailed knowledge of where vessels have been fishing within 40E4. The apparent sudden decline in landings from the Sound of Jura in 2001 is not considered to be associated with a sudden change in fishing practices and is thought more likely to be due to changes in fishery office recording practices. For this reason, the landings split is considered unreliable in recent years and the commercial landings data are now presented for the combined Firth of Clyde and Sound of Jura. Given the relative magnitudes of the fisheries (Clyde likely to be much bigger), the commercial data are likely to be more representative of the Clyde.

## Effort data

In 2015, WGCSE agreed that effort should be reported in kW days, as this is likely to be more informative about changes in the actual fleet effort. Effort shows an overall decreasing trend but was stable through 2010 to 2012 (Figure 17.2.1). Effort increased in 2016 in comparison to 2015, but had been on a generally decreasing trend until 2021, when a slight increase in effort was observed following the extensive interruptions the fishery experienced in 2020 due to the COVID-19 pandemic. Note that the effort time-series range (2000-2021) does not match the more extensive year range available for landings due to a lack of confidence in the reliability of older effort data in the Marine Scotland Science database. The effort is also slightly inconsistent with the landings data because effort is provided for TR2 vessels only, while the 'Nephrops trawl' landings also includes landings by large mesh trawlers targeting Nephrops.

## Sampling levels

Length compositions of landings and discards are obtained during market and on-board observer sampling respectively. These sampling levels are shown in Table 12.2.2 (see. Section 12 FU11 North Minch). Sampling of landings length compositions in the Sound of Jura is more infrequent but samples have been included in the FU13 raising procedure when available. Length compositions for the creel fishery are available for landings only. This is because survival in the animals that are discarded (although little quantitative information exists) is assumed to be high (ICES, 2013). Therefore, these animals are not considered to be removed from the population, and $100 \%$ survival is assumed (ICES, 2013).

## Length compositions

Although assessments based on detailed catch analysis are not presently carried out, examination of length compositions can provide a preliminary indication of exploitation effects. Figure 14.2.2 shows a series of annual Clyde length-frequency distributions for the period 2002 to 2021. Catch and landings length compositions, and mean size are shown for each sex. The mean sizes of both sexes have fluctuated around relatively small ranges since 2015. The mean size of females in the catch has remained relatively stable over the past three to four years, whereas the mean size of males showed an increase until 2020, and then a slight decrease in 2021.

## Sex ratio

Sex ratio in FU 13 shows some variation but males generally make the largest contribution to the annual landings (Figure 14.2.3(a)). This occurs because males are available throughout the year and the fishery takes place in all quarters, although effort is generally reduced during the winter months because of poor weather. Females on the other hand are mainly taken in the summer when they emerge after egg hatching. The seasonal change in proportion of males to females is evident in Figure 14.2.3(b) where males typically dominate in quarters one and four but the ratio is generally more even in quarters two and three. In 2016, males dominated in all quarters, but this was within the observed range of variation typically seen for this stock. The pattern was again fairly typical between 2017 and 2019, but in 2020 all quarterly sex ratios were majority male due to the decreased number of samples which were available for the year. Sampling was also reduced in 2021, but an increased proportion of females was observed in quarter 2. This metric is used as an indicator, whereby increasing proportions of females in the catch might signal an effect of acute overfishing. In this case, however, the atypical sex ratios observed in 2020 (and to some degree 2021) are known to be due to poor sampling, and not a cause for concern to management.

## Mean weights

The mean weights in the landings have fluctuated in this FU over the time-series. Since 2015 there has been a variable but somewhat decreasing trend in this metric. Mean weight for FU13 is generally lower than other areas over the time-series (Table 12.2.3). There is a trend of increasing mean weights in the samples of landings for creel catches, noticeable for both sexes, but particularly for males in the early years of the time-series. However, this has declined in recent years, although sampling levels are low, particularly in the early and most recent years of the timeseries. Given the seasonal variation present in other FUs it is not possible to state with any certainty that this trend is real (Figures 12.2.4 and 12.2.5; see. Section 12 FU11 North Minch).

## Discarding

Discarding of undersized and unwanted Nephrops occurs in the Clyde fishery, and discard sampling has been conducted on the Scottish Nephrops trawler fleet since 1990. Discard rates have been high in this FU and have averaged around $25 \%$ by number in this FU since 1999. Since 2010, discard rates have been estimated to be substantially lower than the long term average, and in 2018 were at the lowest rate in the time-series at only $2.5 \%$ (Table 14.2.2). The discard rate in FU13 increased substantially in 2019 to 19.1\%. Due to an absence of discard sampling in 2020 and 2021, a discard rate of $10.5 \%$ ( $1.8 \% / 26.8 \%$ ) was calculated based on mean rate across all quarters 2017-2019, and allocated to all quarters in 2020 and 2021. The discard rate (adjusted to account for survival) which will be used in the forecast was estimated by taking a three-year average 2019-2021, amounting to $10.4 \%$ ( $8 \% / 14.8 \%$ ).

Studies (Charuau et al., 1982; Sangster et al., 1997; Wileman et al., 1999) suggest that some Nephrops survive the discarding process. An estimate of $25 \%$ survival is assumed for this FU in order to calculate removals (landings + dead discards) from the population. The discard survival rate for creel caught Nephrops has been shown to be high (ICES, 2013) and a value of $100 \%$ is used.

## Abundance indices from UWTV surveys

An underwater TV survey of the stock is conducted annually according to standards set out by the Manual for the Nephrops Underwater TV Surveys (Dobby et al., 2021). Surveys have been carried out in both subareas since 1995 although the Sound of Jura has been surveyed more infrequently. Underwater television surveys of Nephrops burrow distributions avoid the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops. TV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows. Full details of the UWTV approach can be found in the stock annex and the report of WKNEPH in 2009 (ICES, 2009). On average, 37 stations have been considered valid each year for the Firth of Clyde and 11 for the Sound of Jura. These are raised to the estimated ground area available for Nephrops; $2080 \mathrm{~km}^{2}$ based on contoured superficial sediment information (British Geological Surveys).

In 2022, 30 valid stations were used in the final survey analysis for the Firth of Clyde (Table 17.2.3) and 12 stations for the Sound of Jura (Table 14.2.4). Table 14.2 .5 shows a detailed breakdown of information from the most recent TV surveys conducted in the Firth of Clyde. This includes estimates of abundance and variability in each of the strata adopted in the stratified random approach. Details for the Sound of Jura are shown in Table 14.2.6. A CV (coefficient of variation, or relative standard deviation) of $<20 \%$ is considered an acceptable precision level for UWTV survey estimates of abundance (SGNEPS, ICES, 2012). CVs for the three most recent TV surveys in Firth of Clyde and Sound of Jura are lower than the precision level agreed.

Figure 14.2.4 shows the distribution of stations in recent TV surveys (2011-2022) across FU13 (the two distinct subareas can be clearly seen) with the size of the symbols proportional to the Nephrops burrow density. Table 14.2.3 and Figure 14.2.5 show the time-series estimated abundance for the TV surveys in the Firth of Clyde, with $95 \%$ confidence intervals on annual estimates. Similar information for the Sound of Jura is shown in Table 14.2.4 and Figure 17.2.6. Most surveys have detected generally higher densities in the southern part of the Clyde.

The TV survey estimates of abundance for Nephrops in the Firth of Clyde suggest that the population increased until the mid-2000s implying a sustained period of increased recruitment. Following this, abundance has fluctuated around the values previously observed in the early 2000s.

In 2022, the overall abundance increased slightly but remained within recently observed ranges (Figure 14.2.5).

There is not a continuous time-series of abundance in the Sound of Jura and in some years (particularly 2002 and 2006) estimates are associated with large confidence intervals. Abundance has fluctuated with no obvious trend. In 2013, the abundance was at the second lowest point in the time-series. The abundance appeared relatively stable from 2017 until 2021, but in 2022 it dropped below the $15 \%$ quantile of estimated values (Figure 14.2.6).

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009; ICES, 2013). A number of potential biases were highlighted including those due to edge effects, species burrow misidentification and burrow occupancy. The cumulative relative to absolute conversion factor estimated for FU13 was 1.19 meaning that the TV survey is likely to overestimate Nephrops abundance by $19 \%$.

### 14.3 Assessment

## Comparison with previous assessments

The assessment in 2022 is based on a combination of examining trends in fishery indicators and underwater TV survey data, using an extensive data series for the Firth of Clyde component of FU13 and a more limited time-series of UWTV data from the Sound of Jura subarea. The assessment in 2022 follows that of previous years (since 2015) in that the commercial data for Clyde and Sound of Jura have been combined, because of concerns regarding the accuracy of the landings data. There are also no discard samples and limited market samples available for the Sound of Jura. Therefore, the harvest rate and catches for the two areas are presented as a combined total. Nephrops abundance will continue to be monitored separately, with a TV survey being conducted in both subareas where logistically possible.

## State of the stock

The underwater TV surveys are presented as the best available information on the stocks of Nephrops in the two subareas of FU13. The surveys provide fishery-independent estimates of Nephrops abundance. At present, it is not possible to extract any length or age-structure information from the survey and it therefore only provides information on abundance over the area of the survey.

TV survey estimated stock abundance for the Firth of Clyde in 2022 was 1665 million individuals, a $17.7 \%$ increase from the 2021 estimate, well above the $\mathrm{B}_{\text {trigger }}$ value of 580 million. The abundance estimate for the Sound of Jura in 2022 was 241 million individuals, a $22.3 \%$ decrease from the 2021 estimate, but again above the $B_{\text {trigger }}$ value of 160 million.

The harvest rate for the FU13 in 2021 (dead removals for both subareas/ Firth of Clyde and Sound of Jura TV abundance $=21 \%$ ) was above the Fmsy proxy value (the value associated with high long-term yield and low risk depletion) for the Clyde (15.1\%), and the Sound of Jura (12.0\%). Note the Fmsy proxy values for this stock were revised in October 2015 at WKMSYRef4 (ICES, 2016b).

### 14.4 Catch option table

Landings predictions and catch options at various harvest rates (based on principles established at WKNEPH (ICES, 2009)), will be made for Firth of Clyde and Sound of Jura on the basis of the

2022 UWTV survey conducted in June. These will be presented in October 2022 for the provision of advice.

Catch scenario table inputs and historical estimates of mean weight in landings and harvest rates are presented in Table 17.2.2 and summarised below. The calculation of catch options for the Firth of Clyde follows the procedure outlined in the stock annex.

The table below shows the agreed inputs to the catch scenarios table for FU13.

| Input | Data | 2022 assessment |
| :--- | :--- | :---: |
| Survey abundance (millions) | UWTV 2022 | 1665 Clyde; 241 SoJ |
| Mean weight in projected <br> landings (g) | $2019-2021$ | 17.16 |
| Mean weight in projected <br> discards (g) | $2019-2021$ | 7.33 |
| Projected discards | Average 2019-2021 (proportion by number; combined for Firth of <br> Clyde and Sound of Jura) | $13.4 \%$ |
| Discards survival | Proportion by number (assumed) | $25 \%$ |
| Dead projected discards | $2019-2021$ | $10.4 \% *$ |

* Based on mean discard rate (2017-2019) allocated to all Quarters of $2020 \& 2021$. Estimates of $14.8 \%$ and $8 \%$ were derived based on the maximum and minimum observed discard rates, respectively, for the same period


### 14.5 Reference points

Fmsy proxy for this stock was revised in October 2015 at WKMSYRef4 (ICES, 2016a; ICES, 2016b). These were updated on the basis of an average of estimated Fmsy proxy harvest rates over a period of years, which corresponds more closely to the methodology for finfish. In cases where there is a clear trend in the values a five-year average was chosen. Similarly, the five-year average of the F at $95 \%$ of the YPR obtained at the Fmsy proxy reference point was proposed as the Fmsy lower bound and the five year average of the F above $\mathrm{F}_{\mathrm{MAX}}$ that leads to YPR of $95 \%$ of the maximum as the upper bound. Using an average value also has the advantage of reducing the effect of any unusually high or low estimates of the FMSY proxy, which occasionally appear. For this functional unit the FMSY proxy has been revised to $15.1 \%$ for the Clyde and $12.0 \%$ for the Sound of Jura respectively.
For Nephrops stocks, MSY Btrigger has been defined as the lowest stock size from which the abundance has increased and is calculated as 579 million individuals for the Firth of Clyde. The advice from WKMSYRef4 (ICES, 2016b) rounded this value to give an MSY $B_{\text {trigger }}$ of 580 million.

MSY Btrigger was not previously proposed for FU13 (SJ) as there were few points in the survey series (due to missing years). WKMSYRef4 stated that the survey series is now considered to be of sufficient length to allow the Bloss (abundance in 1995) to be proposed as the MSY $B_{\text {trigger. }}$. This results in a value of 160 million (ICES, 2016b). Full details are contained in the stock annex.

These should remain under review by WGCSE and may be revised should improved data become available.

Table 17.2.2 and Figure 17.4.1 show the estimated harvest rates over this period. The harvest rate was calculated from the total dead removals for both subareas divided by the combined abundance for the Firth of Clyde TV survey and the Sound of Jura. This does result in some years
were the harvest rate is not calculable as we do not have a full time-series of TV surveys for the Sound of Jura. The combined harvest rate peaked in 2007 at $43.0 \%$ before declining to around the FmSY level for the Clyde in 2010-2011. The harvest rate has fluctuated since then, and increased in 2021 to $19.3 \%$ from $9.4 \%$ in 2020 . It is unlikely that prior to 2006 , the estimated harvest rates are representative of actual harvest rates due to under-reporting of landings.

### 14.6 Management strategies

Scotland has recently established a network of regional Inshore Fisheries Groups (rIFGs), nonstatutory bodies that aim to improve the management of Scotland's inshore fisheries out to six nautical miles, and to give commercial inshore fishermen a strong voice in wider marine management developments. The rIFGs will contribute to regional policies and initiatives relating to management and conservation of inshore fisheries, including impacts on the marine environment and the maintenance of sustainable fishing communities and measures designed to better conserve and sustainably exploit stocks of shellfish and sea fish (including salmon) in their local waters. Although no IFG proposals specific to the management of Nephrops fisheries have yet been adopted, some of the IFG management plans for the Scottish West Coast include spatial management of Nephrops fisheries and the introduction of creel limits.

A weekend ban on mobile gear was introduced in the Clyde in 1986 under a Scottish Statutory Instrument. Mobile gear is banned in the Inshore Clyde from Friday night to Sunday night, as are vessels greater than 21 m in length.

On the 8th of February 2016, phase 1 of the fisheries management measures for inshore MPAs in Scottish waters came into force (SG, 2016). These measures relate to both NCMPA (Marine (Scotland) Act and the UK Marine and Coastal Access Act) and Special Areas of Conservation (EC Habitats Directives - Council Directive 92/43/EEC) both of which have the aim of conserving biological diversity in Scottish waters and along with other protected sites make up Scotland's MPA network (SG, 2017a). Although not specific to the management of the Nephrops fishery they will influence spatial patterns of fishing for Nephrops where controls on the two main gear types, demersal trawls and creels, are implemented on Nephrops habitat. There are three NCMPAs within the Clyde functional unit. The MPA, which extends onto the main patch of Nephrops habitat, is the South Arran NCMPA, within the Firth of Clyde subarea, where a complete ban on demersal vessels greater than 120 gross tonnage has been implemented. Partial closures (i.e. zoned management) for demersal trawlers smaller than this size and creelers are also in place. For Loch Sween, north of the main habitat area in the Sound of Jura subarea, demersal trawling by vessels is banned. However, for trawlers smaller than 75 gross tonnage, temporal closures are in place over some of the area. For the Upper Loch Fyne and Loch Goil NCMPA, just north of the main habitat area in Firth of Clyde subarea, demersal trawling by vessels greater than 75 gross tones is banned and the activity of vessels below this is zoned. Creeling activity is also zoned (SG, 2016). The areas of the NCMPAs relative to the estimated Nephrops habitat within the Clyde functional unit are presented in Figure 17.6.1.

### 14.7 Quality of assessment and forecast

There are concerns over the accuracy of historical landings and effort data and because of this the final assessment adopted is independent of official statistics. Harvest rates since 2006 are also considered more reliable due to more accurate landings data reported under new legislation.

One of the main issues for this FU is the problem of not being able to split the landings between the Sound of Jura and Firth of Clyde. This means that we are unable to provide harvest rates for the two subareas separately. What is currently provided is not actually a harvest rate for either
sub area; but is likely more representative of the Firth of Clyde. This has an impact on the quality of the assessment but not on the forecast.

In recent years, the length and sex composition of the landings data is considered to be well sampled. However, in 2018 sampling levels fell below this normal standard. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in the Firth of Clyde subarea fishery since 1990, and is considered to represent the fishery adequately. There are few samples available from the Sound of Jura and these have been included in the FU13 raising procedure.

Discard sampling in 2021 was impacted by the COVID-19 pandemic, with no valid samples collected in Functional Unit 13. Estimates of discard rates for all quarters in the assessment were based on mean discard rates across all quarters from 2017-2019 (see "InterCatch", above). This change is considered to have had minimal impact on the quality of the assessment because discard rates have been consistently low in recent years.
Underwater TV surveys have been conducted for this stock every year since 1995. The number of valid stations in the survey has remained relatively stable throughout the time period. Confidence intervals around the abundance estimates are stable throughout the series and relatively low compared with other FUs in area 6.a. In the provision of catch scenarios based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. A three-year average (2019-2021) of discard rate (adjusted to account for some survival of discarded animals) has been used in the calculation of catch advice.

The cumulative relative to absolute conversion factor estimates for FU13 component is largely based on expert opinion (see stock annex). The precision of these bias corrections cannot yet be characterised. The method to derive landings for the catch options is sensitive to the input dead discard rate and mean weight in landings and this introduces uncertainties in the catch forecasts. Precision estimates are needed for these forecast inputs.

The overall area of the ground is estimated from the available BGS contoured sediment data and at present is considered to be a minimum estimate. VMS data, recently made available and linked to landings (from queries of the Scottish FIN database) suggest no major differences between areas fished and the mud sediment maps. The inclusion of vessels smaller than 15 m would likely increase the fished area in some of the inshore locations, while in the Clyde the non-estimated sea loch areas are relatively small.

### 14.8 Recommendation for next benchmark

This stock was last benchmarked in 2009 (ICES, 2009). WGCSE recommends that the issue concerning the split of landings between Sound of Jura and the Firth of Clyde be examined when this stock is next proposed for benchmark process.

### 14.9 Management considerations

ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES division level. Management at the Functional Unit level could provide controls to ensure effort and catch were in line with resources available. In this FU, the two subareas imply that additional controls may be required to ensure that the landings taken in each subarea are in line with the landings advice.

Creel fishing takes place in part of this area although the relative scale of the fishery is smaller than in the Minches. Overall effort in terms of creel numbers is not known, and measures to
control numbers are not in place. There is a need to ensure that the combined effort from all forms of fishing is taken into account when managing this stock.

There is a bycatch of other species in the area of the Firth of Clyde and estimated discards of whiting and haddock by the TR2 fleet are generally high in area 6.a. It is important that efforts continue to ensure that unwanted bycatch is kept to a minimum in this fishery. Current efforts to reduce discards and unwanted bycatches of cod include the implementation of large square meshed panels (SMPs) of 120 mm under the west coast emergency measures, and SMPs of 200 mm implemented as part of the previous Scottish Conservation Credits scheme. A seasonal closure (early spring) in the southwest part of the Firth of Clyde is in place to protect spawning cod although Nephrops vessels are derogated to fish in those parts where mud sediments are distributed.

### 14.10 References

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Table 14.1.1. Nephrops, Clyde and Sound of Jura (FU13), ICES estimates of landings of Nephrops, 1981-2021.

| UK SCOTLAND |  |  |  |  |  | OTHER UK | IRELAND | $\begin{aligned} & \text { TO- } \\ & \text { TAL** } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | NEPHROPS TRAWL | OTHER | CREEL | BELOW MINIMUM SIZE | $\begin{aligned} & \text { SUB TO- } \\ & \text { TAL } \end{aligned}$ |  |  |  |
| 1981 | 2498 | 404 | 66 | 0 | 2968 | 0 | 0 | 2968 |
| 1982 | 2372 | 169 | 79 | 0 | 2620 | 0 | 0 | 2620 |
| 1983 | 3889 | 121 | 52 | 0 | 4062 | 14 | 0 | 4076 |
| 1984 | 3070 | 153 | 77 | 0 | 3300 | 10 | 0 | 3310 |
| 1985 | 3921 | 293 | 65 | 0 | 4279 | 7 | 0 | 4286 |
| 1986 | 4073 | 176 | 79 | 0 | 4328 | 13 | 0 | 4341 |
| 1987 | 2860 | 82 | 64 | 0 | 3006 | 3 | 0 | 3009 |
| 1988 | 3507 | 107 | 43 | 0 | 3657 | 7 | 0 | 3664 |


| UK SCOtLAND |  |  |  |  |  | OTHER UK$16$ | IRELAND $\square$ <br> 0 | то- <br> TAL** <br> 2812 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | NEPHROPS TRAWL | OTHER | CREEL | BELOW MINIMUM sIZE | SUB TO- <br> TAL |  |  |  |
| 1989 | 2577 | 184 | 35 | 0 | 2796 |  |  |  |
| 1990 | 2731 | 121 | 23 | 0 | 2875 | 34 | 0 | 2909 |
| 1991 | 2844 | 145 | 26 | 0 | 3015 | 23 | 0 | 3038 |
| 1992 | 2530 | 247 | 9 | 0 | 2786 | 17 | 0 | 2803 |
| 1993 | 3200 | 110 | 5 | 0 | 3315 | 28 | 0 | 3343 |
| 1994 | 2503 | 50 | 28 | 0 | 2581 | 49 | 0 | 2630 |
| 1995 | 3766 | 131 | 26 | 0 | 3923 | 64 | 0 | 3987 |
| 1996 | 3880 | 108 | 27 | 0 | 4015 | 42 | 0 | 4057 |
| 1997 | 3486 | 46 | 26 | 0 | 3558 | 63 | 0 | 3621 |
| 1998 | 4540 | 79 | 39 | 0 | 4658 | 183 | 0 | 4841 |
| 1999 | 3476 | 29 | 37 | 0 | 3542 | 210 | 0 | 3752 |
| 2000 | 3142 | 63 | 75 | 0 | 3280 | 137 | 0 | 3417 |
| 2001 | 2890 | 65 | 95 | 0 | 3050 | 132 | 0 | 3182 |
| 2002 | 3075 | 53 | 105 | 0 | 3233 | 151 | 0 | 3384 |
| 2003 | 2954 | 20 | 119 | 0 | 3093 | 80 | 0 | 3173 |
| 2004 | 2619 | 8 | 88 | 0 | 2715 | 258 | 0 | 2973 |
| 2005 | 3148 | 5 | 94 | 0 | 3247 | 148 | 0 | 3395 |
| 2006 | 4356 | 1 | 179 | 0 | 4536 | 244 | 0 | 4780 |
| 2007 | 6069 | 4 | 221 | 0 | 6294 | 366 | 0 | 6660 |
| 2008 | 5320 | 3 | 184 | 0 | 5507 | 416 | 0 | 5923 |
| 2009 | 4304 | 1 | 191 | 0 | 4496 | 283 | 0 | 4779 |
| 2010 | 5162 | 5 | 211 | 0 | 5378 | 465 | 0 | 5843 |
| 2011 | 5664 | 9 | 219 | 0 | 5892 | 540 | 0 | 6432 |
| 2012 | 5617 | 4 | 203 | 0 | 5824 | 863 | 0 | 6687 |
| 2013 | 4708 | 4 | 212 | 0 | 4924 | 511 | 0 | 5435 |
| 2014 | 4770 | 1 | 258 | 0 | 5029 | 1178 | 0 | 6207 |
| 2015 | 4035 | 8 | 206 | 0 | 4249 | 898 | 0 | 5147 |


| UK SCOTLAND |  |  |  |  |  | OTHER UK | IRELAND | $\begin{aligned} & \text { TO- } \\ & \text { TAL** } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | NEPHROPS TRAWL | OTHER | CREEL | BELOW MINIMUM SIZE | $\begin{aligned} & \text { SUB TO- } \\ & \text { TAL } \end{aligned}$ |  |  |  |
| 2016 | 4922 | 6 | 267 | 0 | 5195 | 1252 | 4 | 6447 |
| 2017 | 4195 | 3 | 263 | 0 | 4461 | 942 | 1 | 5403 |
| 2018 | 3574 | 13 | 253 | 0 | 3840 | 303 | 0 | 4143 |
| 2019 | 3834 | 3 | 265 | 0 | 4102 | 581 | 0 | 4683 |
| 2020 | 2869 | 10 | 225 | 0 | 3104 | 532 |  | 3636 |
| 2021 | 3805 | 50 | 233 | 0 | 4088 | 907 |  | 4995 |

Table 14.2.1. Nephrops, Clyde (FU13), ICES estimated landings of Nephrops, in each of the subareas (Firth of Clyde and Sound of Jura 1981-2021).

| YEAR | UK LANDINGS |  |  |
| :---: | :---: | :---: | :---: |
|  | FIRTH OF CLYDE | SOUND OF JURA | ALL SUBAREAS |
| 1981 | 2277 | 691 | 2968 |
| 1982 | 1983 | 637 | 2620 |
| 1983 | 3395 | 681 | 4076 |
| 1984 | 2600 | 710 | 3310 |
| 1985 | 3561 | 725 | 4286 |
| 1986 | 3228 | 1113 | 4341 |
| 1987 | 2408 | 601 | 3009 |
| 1988 | 3509 | 155 | 3664 |
| 1989 | 2595 | 217 | 2812 |
| 1990 | 2592 | 317 | 2909 |
| 1991 | 2654 | 384 | 3038 |
| 1992 | 2383 | 420 | 2803 |
| 1993 | 2766 | 577 | 3343 |
| 1994 | 2095 | 535 | 2630 |
| 1995 | 3692 | 295 | 3987 |
| 1996 | 3671 | 386 | 4057 |
| 1997 | 3135 | 486 | 3621 |
| 1998 | 4373 | 468 | 4841 |


| YEAR | UK LANDINGS |  |  |
| :---: | :---: | :---: | :---: |
|  | FIRTH OF CLYDE | SOUND OF JURA | ALL SUBAREAS |
| 1999 | 3423 | 329 | 3752 |
| 2000 | 3229 | 188 | 3417 |
| 2001 | 2979 | 203 | 3182 |
| 2002 | 3350 | 34 | 3384 |
| 2003 | 3154 | 19 | 3173 |
| 2004 | 2965 | 8 | 2973 |
| 2005 | 3388 | 7 | 3395 |
| 2006 | 4768 | 12 | 4780 |
| 2007 | 6580 | 80 | 6660 |
| 2008 | 5845 | 78 | 5923 |
| 2009 | 4688 | 91 | 4779 |
| 2010 | 5782 | 61 | 5843 |
| 2011 | 6363 | 69 | 6432 |
| 2012 | 6634 | 53 | 6687 |
| 2013 | NA | NA | 5435 |
| 2014 | NA | NA | 6207 |
| 2015 | NA | NA | 5147 |
| 2016 | NA | NA | 6447 |
| 2017 | NA | NA | 5403 |
| 2018 | NA | NA | 4143 |
| 2019 | NA | NA | 4683 |
| 2020 | NA | NA | 3636 |
| 2021 | NA | NA | 4995 |

Table 14.2.2. Nephrops, Clyde (FU13): Firth of Clyde and Sound of Jura combined. Adjusted TV survey abundance (Firth of Clyde subarea), landings, discard rate (proportion by number) and estimated harvest rate. The harvest rate was calculated from the total (dead) removals in number for both subareas divided by the combined abundance from both TV surveys.

| YEAR | LANDINGS IN NUMBERS (MILLIONS) | DISCARD <br> IN NUMBERS <br> (MILLIONS) | REMOVALS <br> IN NUMBERS (MILLIONS)** | ADJUSTED SURVEY <br> CLYDE (MILLIONS) | ADJUSTED <br> SURVEY <br> JURA <br> (MIL- <br> LIONS) | COM- <br> BINED <br> HARVEST <br> RATE* | LANDINGS (TONNES) | DISCARDS (TONNES) | DEAD DISCARDS (TONNES) | DIS- <br> CARD RATE <br> (\%) | DEAD DISCARD RATE (\%) | MEAN WEIGHT IN LANDINGS (gr) | MEAN WEIGHT IN DISCARDS (gr) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 207 | 82 | 269 | 579 | 160 | 36.40 | 3987 | 619 | 464 | 28.4 | 22.90 | 19.24 | 7.54 |
| 1996 | 187 | 61 | 233 | 935 | 171 | 21.07 | 4057 | 635 | 476 | 24.7 | 19.70 | 21.68 | 10.35 |
| 1997 | 150 | 70 | 202 | 1198 | NA | NA | 3621 | 598 | 448 | 32 | 26.10 | 24.21 | 8.50 |
| 1998 | 269 | 187 | 409 | 1262 | NA | NA | 4841 | 1292 | 969 | 41 | 34.20 | 17.98 | 6.92 |
| 1999 | 216 | 93 | 286 | 930 | NA | NA | 3752 | 566 | 424 | 30.2 | 24.50 | 17.39 | 6.05 |
| 2000 | 171 | 48 | 207 | 1411 | NA | NA | 3417 | 470 | 352 | 22 | 17.40 | 19.96 | 9.75 |
| 2001 | 164 | 82 | 225 | 1486 | 272 | 12.80 | 3182 | 677 | 508 | 33.5 | 27.40 | 19.46 | 8.23 |
| 2002 | 207 | 50 | 245 | 1571 | 398 | 12.44 | 3384 | 406 | 305 | 19.5 | 15.40 | 16.35 | 8.12 |
| 2003 | 166 | 134 | 266 | 1817 | 260 | 12.81 | 3173 | 1247 | 935 | 44.7 | 37.70 | 19.13 | 9.31 |
| 2004 | 158 | 168 | 284 | 1970 | NA | NA | 2973 | 1435 | 1076 | 51.5 | 44.30 | 18.80 | 8.54 |
| 2005 | 189 | 69 | 241 | 1959 | 303 | 10.65 | 3395 | 611 | 458 | 26.8 | 21.60 | 17.96 | 8.81 |
| 2006 | 248 | 55 | 290 | 1851 | 430 | 12.71 | 4780 | 515 | 386 | 18.2 | 14.30 | 19.27 | 9.31 |
| 2007 | 350 | 387 | 640 | 1233 | 255 | 43.01 | 6660 | 2566 | 1924 | 52.5 | 45.30 | 19.05 | 6.64 |
| 2008 | 357 | 207 | 512 | 1769 | NA | NA | 5923 | 1433 | 1075 | 36.6 | 30.30 | 16.59 | 6.94 |


| YEAR | LANDINGS IN NUMBERS (MILLIONS) | DISCARD <br> IN NUMBERS (MILLIONS) | REMOVALS <br> IN NUMBERS (MILLIONS)** | ADJUSTED SURVEY CLYDE (MILLIONS) | ADJUSTED SURVEY JURA (MILLIONS) | COM- <br> BINED <br> HARVEST <br> RATE* | LANDINGS (TONNES) | DISCARDS (TONNES) | DEAD DISCARDS (TONNES) | DIS- <br> CARD RATE <br> (\%) | DEAD DISCARD RATE (\%) | MEAN WEIGHT IN LANDINGS (gr) | MEAN WEIGHT IN DISCARDS (gr) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 261 | 169 | 388 | 1499 | 251 | 22.17 | 4779 | 1390 | 1043 | 39.3 | 32.70 | 18.31 | 8.23 |
| 2010 | 276 | 55 | 317 | 1750 | 376 | 14.91 | 5843 | 536 | 402 | 16.7 | 13.10 | 21.21 | 9.68 |
| 2011 | 333 | 74 | 388 | 2165 | 312 | 15.66 | 6432 | 568 | 426 | 18.2 | 14.30 | 19.34 | 7.65 |
| 2012 | 306 | 93 | 376 | 1421 | 371 | 20.98 | 6687 | 1066 | 800 | 23.4 | 18.60 | 21.83 | 11.42 |
| 2013 | 262 | 62 | 309 | 1990 | 198 | 14.12 | 5435 | 454 | 341 | 19 | 15.00 | 20.72 | 7.37 |
| 2014 | 295 | 78 | 353 | 1328 | 231 | 22.64 | 6207 | 696 | 522 | 20.9 | 16.60 | 20.79 | 8.92 |
| 2015 | 232 | 54 | 273 | 1820 | 376 | 12.43 | 5147 | 401 | 301 | 18.9 | 14.80 | 22.21 | 7.43 |
| 2016 | 364 | 69 | 416 | 1946 | 422 | 17.57 | 6447 | 636 | 477 | 15.9 | 12.40 | 17.70 | 9.21 |
| 2017 | 316 | 32 | 340 | 1568 | 306 | 18.1 | 5403 | 275 | 199 | 9.5 | 7.1 | 17.02 | 8.55 |
| 2018 | 268 | 7 | 273 | 2193 | 275 | 11.1 | 4143 | 68 | 51 | 2.5 | 1.9 | 16.14 | 9.79 |
| 2019 | 271 | 64 | 319 | 2083 | 318 | 13.3 | 4683 | 435 | 326 | 19.1 | 15 | 17.26 | 6.81 |
| 2020 | 195 | 23 | 212 | 1941 | NA | 9.4 | 3636 | 174 | 130 | 10.5 | 8.1 | 18.96 | 7.59 |
| 2021 | 328 | 38 | 356 | 1414 | 310 | 20.6 | 4995 | 292 | 219 | 10.5 | 8.1 | 15.27 | 7.59 |
| 2022 | - | - | - | 1665 | 241 | - | - | - | - | - | - | - | - |
| Average*** |  |  |  |  |  |  |  |  |  |  | 10.4 | 17.16 | 7.33 |

## * Harvest rates previous to 2006 are unreliable.

** Removals numbers take the dead discard rate into account.
*** Dead discard average: 2019-2021; Mean weight in landings and discard average: 2019-2021.

Table 14.2.3. Nephrops, Clyde (FU13): Firth of Clyde subarea. Results of the 1995-2022 TV surveys (values adjusted for bias).

| YEAR | NUMBER OF VALID STATIONS | MEAN DENSITY (BURROWS / m²) | ABUNDANCE (MILLIONS) | 95\% CONFIDENCE INTERVAL (MILLIONS) |
| :---: | :---: | :---: | :---: | :---: |
| 1995 | 29 | 0.277 | 579 | 176 |
| 1996 | 38 | 0.454 | 935 | 242 |
| 1997 | 31 | 0.571 | 1198 | 262 |
| 1998 | 38 | 0.605 | 1262 | 213 |
| 1999 | 39 | 0.445 | 930 | 289 |
| 2000 | 40 | 0.681 | 1411 | 246 |
| 2001 | 39 | 0.714 | 1486 | 268 |
| 2002 | 36 | 0.756 | 1571 | 288 |
| 2003 | 37 | 0.874 | 1817 | 292 |
| 2004 | 32 | 0.95 | 1970 | 367 |
| 2005 | 44 | 0.941 | 1959 | 287 |
| 2006 | 43 | 0.882 | 1851 | 257 |
| 2007 | 40 | 0.597 | 1233 | 218 |
| 2008 | 38 | 0.849 | 1769 | 291 |
| 2009 | 39 | 0.723 | 1499 | 210 |
| 2010 | 37 | 0.84 | 1750 | 327 |
| 2011 | 40 | 1.041 | 2165 | 305 |
| 2012 | 37 | 0.681 | 1421 | 227 |
| 2013 | 34 | 0.956 | 1990 | 246 |
| 2014 | 35 | 0.639 | 1328 | 237 |
| 2015 | 37 | 0.875 | 1820 | 351 |
| 2016 | 37 | 0.935 | 1946 | 249 |
| 2017 | 38 | 0.754 | 1568 | 239 |
| 2018 | 40 | 1.055 | 2193 | 297 |
| 2019 | 38 | 1.002 | 2083 | 381 |
| 2020 | 28 | 0.933 | 1941 | 297 |
| 2021 | 41 | 0.68 | 1414 | 211 |
| 2022 | 30 | 0.8 | 1665 | 316 |

Table 14.2.4. Nephrops, Clyde (FU13): Sound of Jura subarea. Results of the 1995-2022 TV surveys (values adjusted for bias).

| YEAR | NUMBER OF VALID STATIONS | MEAN DENSITY (BURROWS / m ${ }^{2}$ ) | ABUNDANCE (millions) | 95\% CONFIDENCE INTERVAL (millions) |
| :---: | :---: | :---: | :---: | :---: |
| 1995 | 7 | 0.42 | 160 | 58 |
| 1996 | 10 | 0.45 | 171 | 26 |
| 1997 | no surveys |  |  |  |
| 1998 |  |  |  |  |
| 1999 |  |  |  |  |
| 2000 |  |  |  |  |
| 2001 | 13 | 0.71 | 272 | 76 |
| 2002 | 9 | 1.04 | 398 | 167 |
| 2003 | 12 | 0.68 | 260 | 68 |
| 2004 | no survey |  |  |  |
| 2005 | 11 | 0.79 | 303 | 84 |
| 2006 | 10 | 1.13 | 430 | 134 |
| 2007 | 10 | 0.67 | 255 | 58 |
| 2008 | no survey |  |  |  |
| 2009 | 12 | 0.66 | 251 | 68 |
| 2010 | 12 | 0.98 | 376 | 39 |
| 2011 | 12 | 0.82 | 312 | 73 |
| 2012 | 12 | 0.98 | 371 | 61 |
| 2013 | 9 | 0.52 | 198 | 35 |
| 2014 | 9 | 0.61 | 231 | 90 |
| 2015 | 12 | 0.98 | 376 | 127 |
| 2016 | 12 | 1.11 | 422 | 42 |
| 2017 | 12 | 0.80 | 306 | 71 |
| 2018 | 12 | 0.72 | 275 | 53 |
| 2019 | 12 | 0.832 | 318 | 61 |
| 2020 | no survey |  |  |  |
| 2021 | 12 | 0.812 | 310 | 98 |
| 2022 | 12 | 0.632 | 241 | 71 |

Table 14.2.5. Nephrops, Clyde (FU13): Firth of Clyde subarea. Results by stratum of the 2020-2022 TV surveys. Note that stratification was based on a series of sediment strata (M - Mud, SM - Sandy mud, MS - Muddy sand).
$\left.\begin{array}{lllllllll}\hline \text { STRATUM } & \begin{array}{l}\text { AREA } \\ \left(\mathbf{k m}^{2}\right)\end{array} & \begin{array}{l}\text { NUM- } \\ \text { BER } \\ \text { OF } \\ \text { STA- } \\ \text { TIONS }\end{array} & \begin{array}{l}\text { MEAN } \\ \text { BUR- } \\ \text { ROW } \\ \text { DENSITY } \\ \left(\text { no. } / m^{2}\right)\end{array} & \begin{array}{l}\text { OB- } \\ \text { SERVED } \\ \text { VARI- } \\ \text { ANCE }\end{array} & \begin{array}{l}\text { ABUN- } \\ \text { DANCE } \\ \text { (MIL- } \\ \text { LIONS) }\end{array} & \begin{array}{l}\text { STRATUM } \\ \text { VARI- } \\ \text { ANCE }\end{array} & \begin{array}{l}\text { PRO- } \\ \text { POR- } \\ \text { TION OF } \\ \text { TOTAL } \\ \text { VARI- }\end{array} & \begin{array}{l}\text { PRECI- } \\ \text { SION } \\ \text { LEVEL } \\ \text { (CV) }\end{array} \\ \text { ANCE }\end{array}\right]$

Table 14.2.6. Nephrops, Clyde (FU13): Sound of Jura subarea. Results by stratum of the 2019, 2021, and 2022 TV surveys. Note that stratification was based on a series of sediment strata.

| STRA- <br> TUM | AREA <br> (km2) | NUMBER OF STATIONS | MEAN BURROW DENSITY (no./m2) | OBSERVED VARIANCE | ABUN- <br> DANCE <br> (MILLIONS) | STRA- <br> TUM <br> VARI- <br> ANCE | PROPORTION OF TOTAL VARIANCE | SURVEY <br> PRECISION <br> LEVEL SUR- <br> VEY (CV) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2019 \text { TV }$ <br> survey |  |  |  |  |  |  |  |  |
| M | 90 | 2 | 0.689 | 0.088 | 62 | 357 | 0.389 |  |
| SM | 150 | 4 | 0.878 | 0.023 | 131.8 | 128 | 0.139 |  |
| MS | 142 | 6 | 0.874 | 0.129 | 124.1 | 434 | 0.472 |  |
| Total | 382 | 12 |  |  | 317.9 | 919 | 1 | 0.101 |
| 2021 TV <br> survey |  |  |  |  |  |  |  |  |
| M | 90 | 1 | 0.387 | 0.044 | 34.8 | 355 | 0.148 |  |
| SM | 150 | 5 | 0.845 | 0.254 | 126.8 | 1140 | 0.474 |  |
| MS | 142 | 6 | 1.046 | 0.27 | 148.6 | 909 | 0.378 |  |
| Total | 382 | 12 |  |  | 310.2 | 2404 | 1 | 0.157 |
| 2022 TV <br> survey |  |  |  |  |  |  |  |  |
| M | 90 | 2 | 0.626 | 0.01 | 56.3 | 42 | 0.033 |  |
| SM | 150 | 5 | 0.676 | 0.011 | 101.3 | 47 | 0.037 |  |
| MS | 142 | 5 | 0.59 | 0.287 | 83.8 | 1159 | 0.93 |  |
| Total | 382 | 12 |  |  | 241.4 | 1247 | 1 | 0.162 |

## Landings - International




Figure 14.2.1. Nephrops, Clyde (FU13). Long-term landings and effort.


Figure 14.2.2. Nephrops, Clyde (FU13). Catch length-frequency distribution (dotted) and landings (solid) for Nephrops, 2002-2021. Mean size in catches and landings are represented by solid and dashed orange lines, respectively. Vertical dotted lines are minimum conservation reference size ( 25 mm ) and 35 mm .


Figure 14.2.3. (a) Nephrops, Clyde (FU13). Proportion of landed weight by sex (top), by quarter (bottom) from Scottish trawlers.


Figure 14.2.3. (b) Nephrops, Clyde (FU13), quarterly numeric proportions by sex (2007-2021).


Figure 14.2.4. Nephrops, Clyde (FU13), TV survey station distribution and density (mean burrows/m²) for Firth of Clyde and Sound of Jura subareas, 2011-2022. Sound of Jura located to the east. Shaded green and brown areas represent areas of suitable sediment for Nephrops. Bubbles scaled the same. Red crosses represent zero observations.


Figure 14.2.5. Nephrops, Clyde (FU13): Firth of Clyde subarea. Time-series of revised TV survey abundance estimates (adjusted for bias, solid black line), with 95\% confidence intervals (dotted black lines), 1995-2022. The dashed red line is the rounded $\mathrm{B}_{\text {trigger }}$ value of $\mathbf{5 8 0}$ million individuals.


Fig-
ure 14.2.6. Nephrops, Clyde (FU13): Sound of Jura subarea. Time-series of TV survey abundance estimates (adjusted for bias, solid black line) with 95\% confidence intervals (dotted black lines), 1995-2022. The dashed orange line is the rounded $B_{\text {trigger }}$ value of $\mathbf{1 6 0}$ million individuals.


Figure 14.4.1. Clyde (FU13) Nephrops harvest rate, 1995-2021. The harvest rate is calculated by dead removals (both subareas combined)/TV abundances (both sub-areas combined). The dashed and solid lines are the F Msy $^{\text {proxy }}$ harvest
rate (for the Firth of Clyde 15.1\%) and the time-series of estimated harvest rates, respectively. Harvest rates prior to 2006 are considered unreliable.


Figure 17.6.1. The area of Nephrops habitat (Mud, Muddy Sand and Sandy Mud) within the Clyde functional unit (FU13) relative to the areas of the Nature Conservation MPAs (NCMPAs) which fisheries management measures. Areas where demersal trawling is prohibited, restricted (i.e. vessel size restrictions or seasonal closures) and where creeling is prohibited are displayed. For more detailed information see SG (2016). Geographic Coordinate System: OSGB 1936, Datum: OSGB 1936, Projected Coordinate System: British National Grid. Coastline by Wessel and Smith (2016), MPA sites subsetted from NCMPA (SNH, 2015) and SAC (SNH, 2016) layers, management areas by SG (2017b) and functional units generated from merged ICES rectangles (ICES, 2017). Map and modified layers created using ArcGIS (ESRI, 2014).

## 15 Nephrops in Division 7.a (Irish Sea East, FU14)

### 15.1 Nephrops Subarea 7 general section

## Stock description and management units

A TAC is in place for ICES Area 7 which does not correspond to the assessment units. As Nephrops are limited to muddy habitats the distribution of suitable sediment defines the species distribution and the stocks are therefore assessed as eight separate Functional Units. There are also some smaller catches from areas outside these Functional Units. The ICES statistical rectangles covered by the Functional Units in ICES Area 7 are listed in the table below.

| FU no. | Name | ICES <br> Divisions | ICES Statistical rectangles |
| :---: | :---: | :---: | :---: |
| 14 | Irish Sea East | 7a | 35-38E6; 38E5 |
| 15 | Irish Sea West | 7 a | 35E3, 36E3; 35-37 E4-E5; 38E4 |
| 16 | Porcupine Bank | 7b,c,j,k | 31-35 D5-D6; 32-35 D7-D8 |
| 17 | Aran Grounds | 7b | 34-35 D9-E0 |
| 18* | Northwest Irish Coast | 7b | 36-37 D9; 37E0-E1 |
| 19 | Southeast and southwest Irish Coast | 7a,g,j | 31-33 D9-E0; 31E1; 32E1-E2; 33E2-E3 |
| 20-21 | Labadie, Jones and Cockburn bank | 7g,h | 28 E0-E2; 29 E0-E3; 30E1-E3; 31E2 |
| 22 | Smalls Ground | 7 g | 31-32 E3-E4 |

* Landings from FU18 are reported to other statistical rectangles outside FUs as these are minimal. WGCSE will monitor FU18 landings in case of any fishery developments.

Nephrops Functional Units in Subarea 7 (FU 14-22). The TAC covers all of Subarea 7. (Note: Functional Units in Subarea 6 (FU 11-13) also shown):


## Landings Obligation

From 2019, on the West Coast and around Ireland (FU 11-22), any vessels catching Nephrops had to land all Nephrops. High survival exemptions exist for creel caught Nephrops. De minimis exemptions apply to Nephrops vessels, for Subarea 7 allowing them to discard Nephrops, as long as they made up no more than $5 \%$ of the catch.

Minimum Conservation Reference Size (Minimum landing size)
Under the Landing Obligation, minimum landings sizes have been abolished. Instead a Minimum Conservation Reference Size (MCRS) for each species has been introduced. Unless exempt, Nephrops below the MCRS must be landed and may be sold but cannot go for human consumption. In most cases, the MCRS is the same as old MLS, being 25 mm carapace length (or over 85 mm total length) around Ireland (FUs 16-22); the MCRS is $20 \mathrm{~mm} \mathrm{CL}(>70 \mathrm{~mm} \mathrm{TL}$ ) on the West coast (6.a, FUs 11-13) and the Irish Sea (7a, FUs 14-15).
The MCRS implemented for the Irish Sea at 20 mm CL is less than the rest of the ICES Area 7 (set at $25 \mathrm{~mm} C L$ ) and applies to the Irish and UK fleets. A more restrictive regulation is adopted by the French Producers' Organisations ( 35 mm CL or 115 mm TL ) to all French trawlers.

## Management applicable in 2021 and 2022

The TAC is currently set for the whole Area 7 with a special condition for Porcupine Bank (FU 16). The TAC for 2022 is set at 17,038 tonnes, which is a decrease of $5 \%$ compared to the 18,026 tonnes set for 2021. The TAC area includes a number of Nephrops stocks showing different levels of exploitation. A single TAC covering a number of distinct stocks allows the possibility of unrestricted catches being taken from a heavily exploited stock when advice suggests they should be limited.

Details of all regulations including effort controls in place are provided in the stock annex for all functional units under this subarea.

Fishing opportunities for Norway lobster (Nephrops norvegicus) in Division 27.7, with specific restrictions for FU 16.

|  | Allowances 2021 (tonnes) |  | Allowances 2022 (tonnes) | Change from 2021 |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Division 27.7 | FU 16 | Division 27.7 | FU 16 | Division 27.7 | FU 16 |
| Spain | 993 | 992 | 924 | 846 | $-7 \%$ | $-15 \%$ |
| France | 4,023 | 621 | 3,746 | 530 | $-7 \%$ | $-15 \%$ |
| Ireland | 6,102 | 1,194 | 5,682 | 1,016 | $-7 \%$ | $-15 \%$ |
| Union | 11,118 | 2,807 | 10,352 | 2,392 | $-7 \%$ | $-15 \%$ |
| UK | 6,908 | 483 | 6,686 | 412 | $-3 \%$ | $-15 \%$ |
| TAC | 18,026 |  | 17,038 |  | $-5 \%$ |  |

Sources:
Council Regulation (EU) 2021/92 of 28 January 2021 fixing for 2021 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters
Council Regulation (EU) 2022/515 of 31 March 2022 amending Regulation (EU) 2022/109 fixing for 2022 the fishing opportunities for certain fish stocks and groups of fish stocks applicable in Union waters and for Union fishing vessels in certain non-Union waters

## Landings area 7

Table below gives the summary of reported landings by Functional Unit for ICES Area 7.

| Year | FU 14 - Irish Sea East | FU 15 - <br> Irish Sea <br> West | FU 16 - <br> Porcupine Bank | FU 17 - <br> Aran <br> Grounds | *FU 18 - Ireland Northwest Coast | FU 19 - Ireland Southwest and Southeast coast | FU 20-21 - <br> Labadie, Jones, Cockburn | FU 22 - <br> Smalls <br> Grounds | FUs 20+21+22 <br> - All Celtic Sea <br> FUs combined | Other statistical rectangles Outside FUs | Total Landings ICES Subarea 7 | TAC for 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 961 | 7,296 | 1,744 | 481 |  |  |  |  |  | 249 | 10,731 |  |
| 1979 | 900 | 8,948 | 2,269 | 452 |  |  |  |  |  | 237 | 12,806 |  |
| 1980 | 730 | 4,578 | 2,925 | 442 |  |  |  |  |  | 205 | 8,880 |  |
| 1981 | 829 | 7,249 | 3,381 | 414 |  |  |  |  |  | 382 | 12,255 |  |
| 1982 | 869 | 9,315 | 4,289 | 210 |  |  |  |  |  | 234 | 14,917 |  |
| 1983 | 763 | 9,448 | 3,426 | 131 |  |  |  |  | 3,667 | 174 | 17,609 |  |
| 1984 | 602 | 7,760 | 3,571 | 324 |  |  |  |  | 3,653 | 187 | 16,097 |  |
| 1985 | 498 | 6,901 | 3,919 | 207 |  |  |  |  | 3,599 | 194 | 15,318 |  |
| 1986 | 671 | 9,978 | 2,591 | 147 |  |  |  |  | 2,638 | 113 | 16,138 |  |
| 1987 | 449 | 9,753 | 2,499 | 62 |  |  |  |  | 3,409 | 107 | 16,279 | 24,700 |
| 1988 | 462 | 8,586 | 2,375 | 828 |  |  |  |  | 3,165 | 140 | 15,556 | 24,700 |


| Year | FU 14 - Irish Sea East | FU 15 - <br> Irish Sea <br> West | FU 16 - <br> Porcupine Bank | FU 17 - <br> Aran <br> Grounds | *FU 18 - Ireland Northwest Coast | FU 19 - Ireland Southwest and Southeast coast | FU 20-21 - <br> Labadie, Jones, Cockburn | FU 22 - <br> Smalls <br> Grounds | FUs 20 $\mathbf{+ 2 1 + 2 2}$ <br> - All Celtic Sea <br> FUs combined | Other statistical rectangles Outside FUs | Total Landings ICES Subarea 7 | TAC for 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 401 | 8,128 | 2,115 | 347 |  | 899 |  |  | 4,005 | 134 | 16,029 | 26,000 |
| 1990 | 563 | 8,300 | 1,895 | 519 |  | 754 |  |  | 4,290 | 102 | 16,423 | 26,000 |
| 1991 | 747 | 9,554 | 1,640 | 410 |  | 1077 |  |  | 3,295 | 169 | 16,892 | 26,000 |
| 1992 | 427 | 7,541 | 2,015 | 374 |  | 888 |  |  | 4,165 | 409 | 15,819 | 20,000 |
| 1993 | 515 | 8,102 | 1,857 | 372 | 10 | 905 | 3,466 | 1,182 |  | 455 | 16,864 | 20,000 |
| 1994 | 447 | 7,606 | 2,512 | 729 | 126 | 390 | 4,202 | 941 |  | 570 | 17,523 | 20,000 |
| 1995 | 584 | 7,796 | 2,936 | 867 | 26 | 695 | 3,536 | 1081 |  | 397 | 17,917 | 23,000 |
| 1996 | 475 | 7,247 | 2,230 | 528 | 46 | 888 | 2,822 | 937 |  | 623 | 15,796 | 23,000 |
| 1997 | 566 | 9,971 | 2,409 | 841 | 15 | 756 | 2,038 | 944 |  | 340 | 17,880 | 23,000 |
| 1998 | 388 | 9,128 | 2,155 | 1,410 | 78 | 827 | 1,713 | 835 |  | 514 | 17,048 | 23,000 |
| 1999 | 624 | 10,786 | 2,289 | 1,140 | 16 | 579 | 1,152 | 1,775 |  | 322 | 18,683 | 23,000 |
| 2000 | 567 | 8,370 | 910 | 880 | 9 | 696 | 1,778 | 2,890 |  | 243 | 16,343 | 21,000 |
| 2001 | 532 | 7,441 | 1,222 | 913 | 2 | 815 | 1,833 | 2,938 |  | 368 | 16,064 | 18,900 |


| Year | FU 14 - Irish Sea East | FU 15 - <br> Irish Sea <br> West | FU 16 - <br> Porcupine Bank | FU 17 - <br> Aran <br> Grounds | *FU 18 - Ireland Northwest Coast | FU 19 - Ireland Southwest and Southeast coast | FU 20-21 - <br> Labadie, Jones, Cockburn | FU 22 - <br> Smalls <br> Grounds | FUs 20 $\mathbf{+ 2 1 + 2 2}$ <br> - All Celtic Sea <br> FUs combined | Other statistical rectangles Outside FUs | Total Landings ICES Subarea 7 | TAC for 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 577 | 6,793 | 1,327 | 1,154 | 14 | 1,318 | 2,674 | 1,993 |  | 243 | 16,093 | 17,790 |
| 2003 | 376 | 7,052 | 1,064 | 933 | 16 | 1,239 | 2,953 | 2,065 |  | 186 | 15,884 | 17,790 |
| 2004 | 472 | 7,266 | 1,406 | 525 | 22 | 1,074 | 2,443 | 1,828 |  | 161 | 15,197 | 17,450 |
| 2005 | 570 | 6,529 | 2,197 | 778 | 15 | 712 | 2,469 | 2,533 |  | 180 | 15,983 | 19,544 |
| 2006 | 628 | 7,535 | 2,185 | 637 | 14 | 741 | 2,523 | 1,761 |  | 270 | 16,294 | 21,498 |
| 2007 | 959 | 8,424 | 2,074 | 913 | 3 | 957 | 2,419 | 2,950 |  | 206 | 18,905 | 25,153 |
| 2008 | 726 | 10,482 | 1,000 | 1,057 | 1 | 851 | 2,980 | 3,090 |  | 322 | 20,509 | 25,153 |
| 2009 | 693 | 9,166 | 879 | 626 | 10 | 868 | 3,145 | 2,185 |  | 316 | 17,888 | 24,650 |
| 2010 | 583 | 8,929 | 922 | 939 | 7 | 687 | 1,793 | 2,714 |  | 359 | 16,933 | 22,432 |
| 2011 | 561 | 10,159 | 1,278 | 659 | 13 | 643 | 1,237 | 1,636 |  | 110 | 16,296 | 21,759 |
| 2012 | 531 | 10,527 | 1,258 | 1,246 | 28 | 849 | 1,189 | 2,618 |  | 325 | 18,571 | 21,759 |
| 2013 | 495 | 8,672 | 1,141 | 1,295 | 0 | 794 | 1,387 | 2,257 |  | 194 | 16,235 | 23,605 |
| 2014 | 679 | 8,613 | 1,189 | 766 | 0 | 468 | 1,836 | 2,526 |  | 174 | 16,251 | 20,989 |


| Year | FU 14 - Irish Sea East | FU 15 - <br> Irish Sea <br> West | FU 16 - <br> Porcupine Bank | $\text { FU } 17 \text { - }$ <br> Aran Grounds | *FU 18 - Ireland Northwest Coast | FU 19 - Ireland Southwest and Southeast coast | FU 20-21 - <br> Labadie, Jones, Cockburn | FU 22 - <br> Smalls <br> Grounds | FUs 20+21+22 <br> - All Celtic Sea <br> FUs combined | Other statistical rectangles Outside FUs | Total Landings ICES Subarea 7 | TAC for 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 378 | 8,632 | 1,394 | 370 | 0 | 507 | 2116 | 2,350 |  | 80 | 15,827 | 21,619 |
| 2016 | 237 | 7,327 | 2,154 | 641 | 0 | 590 | 2453 | 3,329 |  | 118 | 16,849 | 23,348 |
| 2017 | 265 | 6,149 | 2,632 | 295 | 0 | 420 | 1,849 | 3,560 |  | 137 | 15,307 | 25,356 |
| 2018 | 263 | 5,756 | 2,751 | 536 | 0 | 238 | 1,803 | 1,974 |  | 200 | 13,521 | 29,091 |
| 2019 | 270 | 7,590 | 2,251 | 167 | 0 | 249 | 2,723 | 2,083 |  | 216 | 15,549 | 19,784 |
| 2020 | 232 | 6115 | 1899 | 222 | 0 | 249 | 413 | 1518 |  | 304 | 10887 | 16815 |
| 2021 | 519 | 6779 | 2476 | 498 | 0 | 415 | 736 | 1616 |  | 346 | 13385 | 18026 |
| Average | 559 | 8144 | 2105 | 620 | 16 | 728 | 2196 | 2073 | 3589 | 257 | 15869 | 22169 |

*Landings from FU18 are reported to other statistical rectangles outside FUs as these are minimal since 2013. WGCSE will monitor FU18 landings in case of any fishery developments.

## Nephrops FU14 section

## Type of assessment in 2022

This stock was inter-benchmarked in September 2015 (ICES, 2015) and the assessment and provision of advice through the use of the UWTV survey data and commercial fishery data follows the process defined by the inter-benchmark process and described in the stock annex (updated at WGCSE 2020). The UWTV survey undertaken in the summer 2022 forms the basis of advice for this stock.

## ICES advice published 29 October 2021

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2017-2019, catches in 2022 should be no more than 835 tonnes.

To ensure that the stock in Functional Unit (FU) 14 is exploited sustainably, management should be implemented at the FU level.

ICES notes the existence of a management plan, developed and adopted by one of the relevant management authorities for Subarea 7. ICES considers this plan to be precautionary when implemented at the FU level."

### 15.2 General

## Stock description and management units

The Irish Sea East Nephrops stock (FU14) is in ICES Subarea 7 and comprises ICES rectangles 38E5, 38E6, 37E6, 36E6, 35E6.

In FU 14 Nephrops are caught on two spatially discrete grounds. Most of the fishery takes place on the main ground located between the West coast of England and Isle of Man, additionally there is also fishing activity in a small inshore ground known as Wigtown Bay.


East Irish Sea fishing grounds: $A=$ Main fishing ground; $B=$ Wigtown bay area. Windfarms represented by red polygons. (Source: ICES, 2015).
Main landing ports: Kilkeel, Portavogie, Ardglass, Whitehaven, Maryport

## Fishery in 2021

The Eastern Irish Sea Nephrops fishery is relatively small compared to other FUs in the TAC area. Landings have been generally declining since 2009 (Table 3.8.2), with an isolated high of 679 tonnes in 2014 and another high of 519 tonnes in 2021. Lows were 237 and 232 tonnes in 2016 and 2020, respectively.

The fishery in FU 14 is dominated by UK vessel activity, on average accounting for $91 \%$ of the reported annual international landings since 2000. In 2021, only UK vessels reported landings from FU 14. Of these UK landings, $64.3 \%$ went into Northern Ireland (down from $80.3 \%$ in 2020), and $35.7 \%$ went into England (up from $19.2 \%$ in 2020). Unlike in 2020, no landings were reported in the Isle of Man.

The main fleets targeting Nephrops include single- and twin-rig otter trawlers operating out of ports in England, Wales, Northern Ireland, and the Republic of Ireland. In 2021, the fleet was split, according to landed weights, by $75.0 \%$ from Northern Irish vessels ( $78.7 \%$ in 2020), $22.3 \%$ from English vessels ( $10.9 \%$ in 2020), $2.7 \%$ from Scottish vessels ( $0 \%$ in 2020) and $0 \%$ from Irish vessels ( $10 \%$ in 2020). The TR2 gear class (otter trawls with $70-99 \mathrm{~mm}$ mesh sizes) accounted for $98.8 \%$ ( $99.3 \%$ in 2020) of the landings, with $1.0 \%$ ( $0.7 \%$ in 2020) of landings being reported from the TR1 gear class (otter trawls with $\geq 100 \mathrm{~mm}$ mesh sizes). The remaining $0.2 \%$ were caught in fixed pots.

A more detailed historical fishery description is provided in the stock annex.

## Information from stakeholders

No additional information was provided.

### 15.3 Data

## InterCatch

Data for 2021 were successfully uploaded to InterCatch prior to the 2022 WG meeting. According to the usual allocation procedure, English landings are allocated to English samples, Northern Irish landings are allocated to Northern Irish samples, and all remaining landings are allocated to pooled English and Northern Irish samples. Due to the impacts of the Covid-19 pandemic, no samples were available for 2020. Landings and discards length-frequency distributions for 2017 - 2019 were therefore used for the 2021 assessment. For the 2021 data year, Northern Irish samples were available and were allocated to all landings. Mean weights and discard rates for 2021 were calculated as the average of the values for 2019 and 2021.

## Landings

Official landings as reported to ICES from FU14 are presented in Tables 3.8.1 and 3.8.2. There are reported landings for this functional unit since 1973 with a minimum and maximum of 178.7 t (in 1974) and 960.5 t (in 1978) respectively. Between 1987 and 2006 landings from FU 14 appeared relatively stable fluctuating around a long-term average of about 550 t . The introduction of the UK Buyers and Sellers legislation in 2006 precludes direct comparison with previous years, as reported levels are considered to have significantly improved. Over the period 2007-2020, landings have declined considerably from the peak year of 2007 ( 959 t ); landings in 2020 were the lowest in the period ( 232 t ). There were no reported discards in 2020 and discarding ( 15 t ) has been estimated based on 2017-2019 rates. In 2021, landings increased again to 519 tonnes, all into the UK, with 29 tonnes of reported discards.

## Effort

Following discussions at WGCSE, it was concluded that effort should be reported in the WGCSE report in kWdays, and LPUE should be reported in $\mathrm{kg} / \mathrm{kW}$ days in the knowledge that the trend is likely to be a biased underestimate because it is not adjusted for efficiency or behavioural changes. Effort calculations are likely to have been unreliable prior the introduction of the Buyers and Sellers legislation in 2006 by the UK and improvement in landings reporting.

Total UK and Irish targeted effort ( $70-99 \mathrm{~mm}$ mesh with $>30 \%$ Nephrops by weight) and LPUE is reported in Table 3.8.3 and shown in Figure 3.8.2. Until 2020, there was a general decline in targeted effort since 2007 (although 2014 did see a small rise in effort compared to 2013). In 2021, targeted effort increased again, almost tripling compared to 2020.

Within the UK targeted metier, there are significant differences between sub-fleets and changes in fleet composition may therefore unduly influence catch rate metrics. These issues need to be examined further.

## Sampling Levels

Sampling levels, data aggregating and raising procedures were reviewed by IBPNeph 2015, documented in the stock annex, and examined further at WGCSE 2018. Recent sampling levels have fluctuated; prior to 2016 sample data have only been available from landings into England, however since 2016 samples have also been available from landings into Northern Ireland. In 2020, there was no sampling activity from the fishery due to impact of the Covid-19 pandemic. For 2021, only Northern Irish samples are available.


Number of observer trips on English and Northern Irish vessels, as well as the number of shore-based catch samples.

## Commercial length-frequency distributions

The raised catch length distributions are shown in Figure 3.8.3. Mean individuals weights and carapace lengths are listed in Table 3.8.4. The mean sizes for both sexes from 2008 fluctuate considerably. For 2020, the mean individual weights and carapace lengths of Nephrops were estimated from the average of 2017-2019 values to be slightly lower than in 2019 and 2020 but higher than the record low of 2016. On the basis of Northern Irish samples, the mean weight and carapace length in landings decreased compared with the 2017-2019 average, as well as with the values in 2019, while mean weight and carapace length in discards increased compared with the previous years.

## Length composition

Between 2010 and 2012, sampling levels are considered insufficient to reliably characterise the length composition of extractions. Increased sampling levels from 2013 onwards have allowed for length compositions to be constructed. For 2013 and 2014, a full revision was done through an inter-benchmark process (ICES, 2015; described in the stock annex). Data aggregating and raising procedures from 2015 to 2017 were conducted according to benchmark procedures (ICES, 2005) and referred in the stock annex. These were revised during WGCSE 2018 to account for Northern Irish sampling data since 2016 and are described further in the stock annex. No sampling activity was possible during 2020 due to the effects of COVID-19 pandemic and length composition data from 2017 to 2019 were again used to generate mean sizes. In 2021, the only samples were from Northern Ireland, which were allocated to all landings.

## Sex ratio

Mature females are mainly caught in the non-berried state between moulting (which peaks in May), and spawning (which peaks in September). Females mature at about 23 mm carapace length. (Thomas and José Figueiredo, 1965).

The catch sex ratio by year is shown in Figure 3.8.5. The ratio is quite variable but average sex ratio is $56 \%$ male (1999-2019), the sex ratio for 2019 being slightly above this (62\%). Sex ratio was not estimated in 2020. In 2021, the male sex ratio increased to $73 \%$.

## Mean weight explorations

The annual mean weight estimate for landings and discards is provided in Table 3.8.4 and in Figure 3.8.6. There is a substantial difference between the mean weights prior to 2011 and after 2013 (the gap being where sampling was too low to be reliable). Since 2016, NI sampling has been included and the mean weight of NI samples is considerably lower than for English sampling (e.g. for 2017-2019, mean weight of landings from English sampling was 30.9 g compared to 15.2 g in Northern Irish sampling). As a result, comparison with years prior to 2016 is not practical. Mean weights over the years 2016-2019 have been variable without trend. Mean weights were not estimated for 2020 due to no sampling. Northern Irish sampling resumed in 2021, but due to the small sizes in the Northern Irish samples, compared with the English samples, the mean individual weights and sizes for that year need to be treated with scepticism.

## Discarding

Discard selection was revised at the IBP process in 2015 (ICES, 2015) and described in the stock annex. Figure 3.8.4 shows a single discard ogive fitted by pooling all years (2003-2014) and mesh sizes. Final discard selection for the East Irish Sea shows a L50 $=23.54$ and a L25=24.77 mm CL (Figure 4.3.4), which shows a selectivity at higher sizes compared with FU15. Due to high interannual variation in mean sizes of both landings and discards, the discard ogive was not updated using later data.

Table 3.8.5 gives raised international landings and discard weight and numbers by year.
At IBPNeph (ICES, 2015), it was agreed that the discard survival rate should be updated from $0 \%$ to $10 \%$. Although there are no direct survivability studies available for this area, it is expected that the survivability of discarded animals should be similar to the fishery in FU15 where fishing practices are similar and both are largely spring/summer fisheries and animals discarded are exposed to warmer temperatures before being returned to the sea.

## Abundance indices from UWTV surveys

Since August of 2007, the UK has carried out an underwater TV survey of the Nephrops grounds in the eastern Irish Sea. The survey is of a fixed grid design and is carried out using the same protocols used in UWTV surveys in the western Irish Sea (ICES, 2007; ICES, 2014, Dobby H., et al, 2021). The survey stations used in 2022 are presented in Figure 3.8.7.

Due to the construction of the windfarm in the southern part of the ground the survey area was reviewed at IBP 2015, but the protocols and standardised process to run the survey were not modified (see stock annex and IBP 2015 report ICES, 2015). The new survey area (based on a cokriging model) is shown in Figure 3.8.8. The boundary used to define the ground limits for absolute abundance runs close to the outer survey stations.

| Ground | Area Km $^{\mathbf{2}}$ | Source |
| :--- | :---: | :--- |
| Main ground 2008-2010 | 1032.75 | WGCSE 2008 |
| Main ground 2011-2019 | 1019.79 | IBP 2015 - ICES, 2015 |
| Wigtown Bay | 67.21 | IBP 2015 - ICES, 2015 |

Wigtown Bay in relation to Main ground $=6.6 \%$ * (increase from $1.9 \%$ prior to the windfarm construction).

Abundance indexes were revised back to 2011, the year where the effect of effort displacement is clearly visible due to the windfarm construction. Final updated burrow density estimates are presented in Table 3.8.6 and visualised in Figure 3.8.9 where the geo-spatial model was updated using the new area based on the co-kriging approach ( $1019.79 \mathrm{~km}^{2}$ ) and the extrapolation to Wigtown Bay using 6.6\%.

The abundance estimate for 2022 ( 386 million) is a decrease of $1.7 \%$ compared to the 2021 figure of 393 million (Figure 3.8.10) and 18.9 \% lower than the 2008-2021 average of 476 million. The surveys show a clear spatial distribution pattern, with highest densities in the centre of the patch and more variable in the areas further north and south. The grounds are fairly well delineated by consistently low-density ground to the west (Figure 3.8.9). CVs over the entire time-series (Table 3.8.6) are within the accepted precision level of 20\% (ICES, 2012).

The use of the UWTV surveys for the provision of Nephrops management advice was extensively reviewed by WKNEPH (2009). A number of potential factors were highlighted including those due to edge effects; species burrow misidentification and burrow occupancy. Using the same process adopted at WKNEPH, a cumulative absolute conversion factor for this FU was predicted to be 1.2 for FU 14 (see stock annex) which means the TV survey is likely to overestimate Nephrops abundance by $20 \%$. The burrow abundances shown in Table 3.8.5 and Figure 3.8.9 have been adjusted using this conversion factor since 2008.

In 2021, a new survey camera system was implemented using high-resolution stills-based footage, generally resulting in improved picture and burrow definition. Comparison of the old and new systems in other survey areas (FU 16 and FU 20-21 combined) has shown no significant difference in density estimates and previous assumptions relating to correction factors are still applied.

### 15.4 Assessment

## Comparison with previous assessments

The methods normally used are in line with WKNEPH (ICES, 2009) and the approach taken by WGCSE for other Nephrops stocks in Subareas 27.6 and 27.7. This approach was inter-benchmarked at IBPNeph (ICES, 2015) and is described in the stock annex. Deviations from the standard procedure were required due to the lack of sampling data in 2020. As a result, WGCSE 2021 carried out the assessment for this stock using the same three-year average (2017-2019) for weights and discarding rates as for the 2020 assessment. For this current assessment, the average weights and discarding rates are based on 2019 and 2021 sampling data. Throughout the pandemic, the UWTV surveys were carried out as usual. Therefore, the current abundance values were used for the assessments in 2021 and 2022, according to the established procedure.

## State of the stock

UWTV abundance estimates suggest that the stock size has fluctuated between abundance values of 350 and 694 million Nephrops. The 2022 estimate ( 386 million) decreased by $1.7 \%$ in relation to 2021 but is still above the MSY Btrigger ( 350 million).

Table 3.8 .5 and Figure 3.8 .11 summarise the abundance estimated including the confidence intervals and the harvest ratios (\% total removed / UWTV abundance) which have been above the Fmsy proxy.

### 15.5 Catch scenarios table

Catch scenarios table inputs and historical estimates of mean weight in landings and harvest ratios are presented in Table 3.8.5 and summarised below. The calculation of catch options for FU14 follows the procedure outlined in the stock annex. The basis for the catch options:

| Variable | Value | Notes |
| :--- | :---: | :--- |
| Stock abundance (2023) | 386 | UWTV survey 2022; individuals in millions |
| Mean weight in projected landings | 19.9 | Average 2019-2021; in grammes |
| Mean weight in projected discards | 9.13 | Average 2019-2021; in grammes |
| Projected discard rate | 13.9 | Average 2019-2021; percentage by number of the total <br> catch |
| Discards survival rate* | 10 | Percentage by number of the discards |
| Dead projected discard rate | 12.7 | Average 2019-2021; percentage by number of the total <br> catch |

*Only applied in scenarios where discarding is allowed.

### 15.6 Reference points

Reference points were defined for this stock at the IBPNeph (ICES, 2015) and proposals for Fmsy ranges made by WKMSYRef4 (ICES, 2016a; 2016b).

Based on the fact that some biological parameters are poorly known; inconsistent biological sampling; uncertainties about the stability of the stock over the reference period and uncertainties about the variability of recruitment it is expected that a combined sex $\mathrm{F}_{0.1}$ is a suitable Fmsy proxy for this stock. This corresponds to a harvest rate of $11 \%$ and this value is expected to deliver high long-term yield with a low probability of recruitment over-fishing. These calculations assume that the UWTV survey has knife-edge selectivity at 17 mm and that the supplied length frequencies represented the population in equilibrium. Landings, and correspondingly harvest rates, were significantly higher in 2021 than in previous years. The harvest rate in 2021 was $6.14 \%$, with an average of $2.92 \%$ over the previous three years. Historically the available data show a maximum harvest rate of $8.2 \%$ in 2008 which is below the FMSY proxy.

At the IBPNeph, a MSY Btrigger was defined for this stock. According with this definition, Btrigger was set for FU14 as 350 million, corresponding to the lowest observed abundance estimate from the UWTV time-series, which occurred in 2009.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY Btrigger | 350 million individuals | The lowest observed abundance estimate from the UWTV survey time-series. | ICES (2015) |
|  | FmsY | 11\% harvest rate | Fmsy proxy equivalent to $\mathrm{F}_{0.1}$ for combined sexes. | ICES (2015) |
|  | Fmsy lower | 9.1\% harvest rate | Average of the F at $95 \%$ of the YPR obtained at the $\mathrm{F}_{\text {MSY }}$ proxy reference point | $\begin{aligned} & \text { ICES } \\ & \text { (2016b) } \end{aligned}$ |
|  | Fmsy upper | 11\% harvest rate | Average of the F above $\mathrm{F}_{\text {MAX }}$ that leads to YPR of $95 \%$ of the maximum; capped at $\mathrm{F}_{\mathrm{MSY}}$ | $\begin{aligned} & \text { ICES } \\ & \text { (2016b) } \end{aligned}$ |

### 15.7 Management strategies

There are no explicit management strategies for this stock.
The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to Norway lobster (Nephrops norvegicus) by functional unit in ICES Subarea 7 and also demersal stocks. The plan specifies conditions for setting fishing opportunities, depending on stock status and making use of the FmSY ranges.

### 15.8 Quality of assessment and forecast

The quality of landings data has improved since 2012, but concerns over the accuracy of earlier years limits the period we can be confident about regarding trends in LPUE and landings.

Underwater TV surveys have been conducted annually for this stock since 2007. The quality of the data from the first survey and the limited number of valid stations in the survey limits the number of useable surveys to 2008-2022.

In 2021, the survey camera system and reviewing method changed. Previous assumptions relating to correction factors are still applied. Comparison of the old and new systems in Functional Unit 16 has shown no significant difference in density estimates.

The revised algorithm used to derive distance covered by the sledge is considered significantly more robust than the previous algorithm.

The IBP 2015 managed to address key points:

- Revisions to the area of the Nephrops grounds based on new available data: VMS, UWTV data and sediment information.
- A review of fishery data and raising procedures.
- Review of Reference points: Fmsy proxies and MSY Btrigger.

After this revision the quality of the assessment improved. Although there are still specific uncertainties and assumptions that need to be examined further for the East Irish Sea before less conservative $\mathrm{F}_{\text {msy }}$ proxies could be considered.

There are several key uncertainties and bias sources in the method proposed (these are discussed further in ICES, 2009a). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (ICES, 2007; ICES, 2008; ICES, 2009b). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate but no more precise (ICES, 2009a).

The cumulative absolute conversion factor estimates for FU14 are largely based on expert opinion. However, these were based on experience on other grounds and relatively limited experience on these grounds which would make this less reliable. The precision of these cannot yet be characterised. Ultimately there still remains a degree of subjectivity in the production of UWTV abundance estimates.

The effect of this assumption on realised harvest rates has not been investigated but remains a key uncertainty.

### 15.9 Recommendation for next benchmark

This stock was last benchmarked by IBPNeph (ICES, 2015). WGCSE will keep the stock under close review and recommend future benchmark as required.
At IBP 2015, it was mentioned that there are specific uncertainties and assumptions that need to be examined further for the East Irish Sea before less conservative Fmsy proxies could be considered.

- More accurate mapping of the spatial extent of the grounds and fisheries, this includes having positional data for $<12$ metre vessels and more survey data in Wigtown Bay area to better define this ground. Station grid was extended to Wigtown Bay in 2016.
- For now the total abundance estimate for FU14 is based on the abundance estimates of the geospatial model for the main ground plus adding the area of Wigtown Bay. As this area is becoming a more significant fishing patch it is worth to consider the use of a separate geospatial model in this ground. This should be explored in a future benchmark work.
- Improvement of spatial coverage and sampling of landings and discards, this includes increasing the sampling levels to cover Northern Irish vessels. Northern Irish sampling has been included in the assessment since 2018
- Area specific length-weight and maturity data to validate the parameters used for this FU.
- Better knowledge of the difference in growth and population structure across the area.
- If following the current advice, the recommended catches are taken, then the stock may decrease to well below MSY $B_{\text {trigger }}$ in the short term. The basis for setting MSY $B_{\text {trigger }}$ is currently from recent history may be too high, it could also be due to recent low recruitment (transitory issue) or that the FMSY is too high. As such, the MSY trigger reference point needs to be looked into. It was noted that the basis for MSY $B_{\text {trigger }}$ was the recent history and that the value may be too high.
- Advice is compiled for ADGNEPH in October. Lagged (one year) TV survey gives good correlation with LPUE, could this be used to calculate harvest rate rather than the in-year ratio?


### 15.10 Management considerations

ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES division level. Management at the Functional Unit level could allow effort and catch to be controlled in line with the scale of the resource.
There are no explicit recruitment indices.
The UWTV survey data allow for the provision of catch options and also to adopt the MSY approach. The UWTV surveys are conducted annually and a benchmark process has been adopted
in 2015. In the past this stock has only been assessed biannually. These data provide the opportunity to reassess this stock more reliably on an annual basis.

### 15.11 References

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Table 3.8.1. Irish Sea: Landings (tonnes) by FU. In 2012 and 2013 landings outside FU for Area 7a were not provided, so have been calculated from ICES official landings for 7a minus the FU areas.

| YEAR | FU14 |  | OTHER |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 |  | 567 | 8370 | 1 | 8938 |
| 2001 |  | 532 | 7441 | 3 | 7976 |
| 2002 |  | 577 | 6793 | 1 | 7371 |
| 2003 |  | 376 | 7052 | 3 | 7431 |
| 2004 |  | 472 | 7267 | 25 | 7764 |
| 2005 |  | 570 | 6554 | 103 | 7227 |
| 2006 |  | 628 | 7561 | 52 | 8241 |
| 2007 |  | 959 | 8491 | 83 | 9533 |
| 2008 |  | 676 | 1050 | 122 | 11306 |
| 2009 |  | 708 | 9198 | 57 | 9963 |
| 2010 |  | 582 | 8963 | 23 | 9568 |
| 2011 |  | 561 | 10162 | 61 | 10784 |
| 2012 |  | 531 | 10527 | 208 | 11266 |
| 2013 |  | 495 | 8672 | 89 | 9256 |
| 2014 |  | 679 | 8613 | NA | 9292 |
| 2015 |  | 378 | 8632 | NA | 9010 |
| 2016 |  | 237 | 7327 | 9 | 7564 |
| 2017 |  | 265 | 6149 | 0 | 6414 |
| 2018 |  | 268 | 5756 | 0 | 6024 |
| 2019 |  | 270 | 7590 | 4 | 7864 |
| 2020 |  | 232 | 6115 | 7 | 6354 |
| 2021 |  | 519 | 6779 | 20 | 7318 |

Table 3.8.2. Irish Sea East (FU14): Landings (tonnes) by country of landing and total discards, 2000-2019.


[^7]Table 3.8.3. Irish Sea East (FU14): Effort data for the UK and Irish trawl Nephrops directed fleet.

| YEAR | UK direct fleet |  |  | Irish direct fleet |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | EFFORT (KW DAYS) | LANDINGS (TONNES) | LPUE | EFFORT (KW DAYS) | LANDINGS (TONNES) | LPUE |
| 2006 | 343,249 | 577.2 | 1.7 | 6,932 | 18.3 | 2.8 |
| 2007 | 443,319 | 854.4 | 1.9 | 25,309 | 79.2 | 3.1 |
| 2008 | 366,696 | 628.9 | 1.7 | 8,136 | 14.9 | 1.9 |
| 2009 | 354,210 | 680.1 | 1.9 | 5,516 | 13.1 | 2.5 |
| 2010 | 296,097 | 527.3 | 1.8 | 13,496 | 44.6 | 3.3 |
| 2011 | 252,607 | 525.7 | 2.1 | 8,955 | 29.7 | 3.6 |
| 2012 | 215,851 | 452.4 | 2.1 | 21,224 | 52.8 | 2.6 |
| 2013 | 210,108 | 445.1 | 2.1 | 11,304 | 35.5 | 3.1 |
| 2014 | 279,606 | 636.8 | 2.3 | 10,259 | 28.5 | 2.8 |
| 2015 | 132,751 | 275.7 | 2.1 | 27,128 | 83.7 | 3.1 |
| 2016 | 109,449 | 214.9 | 2.0 | 9,496 | 21.2 | 2.2 |
| 2017 | 101,657 | 252.4 | 2.5 | 2,620 | 6.7 | 2.6 |
| 2018 | 113,740 | 245.8 | 2.2 | 3,042 | 5.2 | 1.7 |
| 2019 | 94,606 | 248.1 | 2.6 | 3,591 | 8.7 | 2.4 |
| 2020 | 61,747 | 203.1 | 3.3 | 7,660 | 22.9 | 3.0 |
| 2021 | 177,961 | 504.8 | 2.8 | - | 0 | - |

Table 3.8.4. Irish Sea East (FU14): Mean size (CL) and weight combined by sex for total annual landings and discards and proportion discarded.

| Year | Mean CL (mm) Landings | Mean CL (mm) Discards | Mean Weight (g) Landings | Mean Weight (g) Discards | Proportion discarded |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 29.83 | 22.32 | 19.05 | 7.52 | 0.26 |
| 2001 | 30.59 | 22.74 | 20.87 | 7.97 | 0.17 |
| 2002 | 30.64 | 23.75 | 22.41 | 8.98 | 0.15 |
| 2003 | 33.69 | 22.43 | 29.12 | 7.62 | 0.10 |
| 2004 | 31.01 | 22.24 | 21.93 | 7.57 | 0.15 |
| 2005 | 30.74 | 23.16 | 21.48 | 8.44 | 0.13 |
| 2006 | 32.36 | 22.75 | 25.07 | 7.98 | 0.10 |
| 2007 | 31.81 | 21.92 | 23.94 | 7.33 | 0.14 |
| 2008 | 31.07 | 23.14 | 22.88 | 8.49 | 0.13 |
| 2009 | 35.57 | 23.21 | 36.49 | 8.58 | 0.04 |

2010*

2011*
2012*

| 2013 | 30.14 | 22.43 | 19.94 | 7.87 | 0.16 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | 31.01 | 24.34 | 22.37 | 9.60 | 0.11 |
| 2015 | 32.05 | 22.57 | 25.19 | 7.82 | 0.13 |
| 2016 ** | 27.39 | 23.11 | 15.82 | 8.38 | 0.14 |
| 2017 | 29.05 | 24.07 | 18.97 | 9.50 | 0.18 |
| 2018 | 30.58 | 24.46 | 21.39 | 9.78 | 0.07 |
| 2019 | 29.49 | 22.90 | 20.93 | 8.40 | 0.12 |
| 2020 *** | 29.71 | 23.81 | 20.43 | 9.23 | 0.12 |
| 2021 **** | 29.35 | 24.09 | 19.56 | 9.07 | 0.17 |

[^8]Table 3.8.5. Irish Sea East (FU14): Sumary table for forecast inputs and historical estimates of raised landings and discards, mean weight in landings and harvest rate.

|  |  |  |  |  |  | UWTV abundance estimate |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
| 2000 | 30 | 11 | 40 | 24.4 | 26.4 |  |  |  | 567 | 80 | 19.05 | 7.52 |
| 2001 | 26 | 5 | 31 | 15.5 | 17.0 |  |  |  | 532 | 42 | 20.87 | 7.97 |
| 2002 | 26 | 5 | 30 | 14.1 | 15.4 |  |  |  | 577 | 42 | 22.41 | 8.98 |
| 2003 | 13 | 1 | 14 | 9.0 | 9.9 |  |  |  | 376 | 11 | 29.39 | 7.64 |
| 2004 | 22 | 4 | 25 | 13.5 | 14.8 |  |  |  | 472 | 28 | 21.93 | 7.57 |
| 2005 | 275 | 4 | 30 | 11.8 | 13.0 |  |  |  | 570 | 33 | 21.48 | 8.44 |
| 2006 | 25 | 3 | 28 | 9.2 | 10.1 |  |  |  | 628 | 22 | 25.07 | 7.98 |
| 2007 | 40 | 6 | 46 | 12.5 | 13.8 |  |  |  | 959 | 47 | 23.94 | 7.33 |
| 2008 | 30 | 4 | 34 | 11.6 | 12.7 | 408 | 63 | 8.2 | 676 | 37 | 22.88 | 8.49 |
| 2009 | 19 | 1 | 20 | 3.3 | 3.7 | 350 | 76 | 5.7 | 707 | 6 | 36.49 | 8.58 |
| 2010 |  |  |  |  |  | 422 | 103 |  | 582 |  |  |  |
| 2011 |  |  |  |  |  | 449 | 99 |  | 561 |  |  |  |
| 2012 |  |  |  |  |  | 694 | 99 |  | 531 |  |  |  |
| 2013 | 25 | 5 | 30 | 15.0 | 16.4 | 487 | 82 | 6.0 | 495 | 39 | 19.94 | 7.87 |
| 2014 | 30 | 4 | 34 | 9.8 | 10.8 | 449 | 92 | 7.5 | 679 | 32 | 22.37 | 9.60 |
| 2015 | 15 | 2 | 17 | 11.9 | 13.0 | 591 | 86 | 2.9 | 378 | 18 | 25.19 | 7.82 |
| 2016* | 15 | 2 | 17 | 12.4 | 13.6 | 430 | 106 | 4.0 | 237 | 20 | 15.82 | 8.38 |
| 2017 | 14 | 3 | 17 | 16.2 | 17.6 | 580 | 89 | 2.9 | 265 | 29 | 18.97 | 9.50 |
| 2018 | 12 | 1 | 13 | 6.3 | 6.9 | 514 | 118 | 2.6 | 263 | 9 | 21.39 | 9.78 |
| 2019 | 13 | 2 | 14 | 11.1 | 12.2 | 399 | 69 | 3.6 | 270 | 15 | 20.93 | 8.40 |
| 2020 ** | 11 | 2 | 13 | 11.2 | 12.3 | 496 | 84 | 2.6 | 232 | 15 | 20.43 | 9.23 |
| 2021 | 21 | 4 | 24 | 13.5 | 15.0 | 393 | 78 | 6.1 | 519 | 58 | 19.56 | 9.07 |
| 2022 |  |  |  |  |  | 386 | 110 |  |  |  |  |  |

Note: Abundance is adjusted by using a cumulative absolute conversion factor of 1.2. Abundance (millions) including Wigtown Bay (1.9\% 2008-2010; 6.6\% 2011-2019). Due to poor sampling no estimates for 2010-2012.

* Values for 2016 revised at WGCSE 2018 due to inclusion of Northern Irish sampling in 2016 and 2017.
** Removals for 2020 calculated using 2020 landings and unweighted average of mean weights from 2017-2019.

Table 3.8.6. Nephrops, Irish Sea East (FU14): Results of the 2008-2020 TV surveys (values adjusted for bias).

| Year | No valid stations | Mean Kriged density (no./m²) | Abundance (millions) including Wigtown Bay (1.9\% 2008-2010) | Abundance (millions) including Wigtown Bay (6.6\% 2011-2018) | $\begin{aligned} & 95 \% \\ & \text { CI } \end{aligned}$ | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 |  |  | Unreliable data |  |  |  |
| 2008 | 32 | 0.38 | 408 |  | 63 |  |
| 2009 | 32 | 0.33 | 350 |  | 76 |  |
| 2010 | 26 | 0.4 | 422 |  | 103 |  |
| 2011 | 26 | 0.41 |  | 449 | 99 | 11.2\% |
| 2012 | 26 | 0.64 |  | 694 | 99 | 7.3\% |
| 2013 | 31 | 0.45 |  | 487 | 82 | 8.5\% |
| 2014 | 34 | 0.41 |  | 449 | 92 | 10.4\% |
| 2015 | 42 | 0.54 |  | 591 | 86 | 7.4\% |
| 2016 | 48 | 0.40 |  | 430 | 106 | 12.6\% |
| 2017 | 45 | 0.53 |  | 580 | 89 | 7.8\% |
| 2018 | 46 | 0.47 |  | 514 | 118 | 11.7\% |
| 2019 | 41 | 0.37 |  | 399 | 69 | 9.3\% |
| 2020 | 43 | 0.46 |  | 496 | 84 | 8.6\% |
| 2021 | 44 | 0.36 |  | 393 | 78 | 10.1\% |
| 2022 | 46 | $0.38$ |  | 386 | 110 | 14.6\% |

Note: Abundance is adjusted by using a cumulative absolute conversion factor of 1.2. Abundance (millions) including Wigtown Bay (1.9\% 2008-2010; 6.6\% 2011-2020).


Figure 3.8.1. Irish Sea East (FU14): Landings in tonnes by country. GBE=England; GBN=Northern Ireland; GBS=Scotland; Rep. of Ireland=Republic of Ireland.

## Directed Effort



Figure 3.8.2. Irish Sea East (FU14): Effort data (KW days) for UK and Irish directed Nephrops fleet.


Figure 3.8.3. Irish Sea East (FU14): Length distribution of landings (solid lines) and catch (dotted lines), 2000-2021. Length frequencies for 2010-2012 are based in very poor sampling so not reliable. No sampling was carried out in 2020. Figure shows a vertical display of MLS $(\mathbf{2 0 ~ m m ~ C L})$ and 35 mm CL levels.

## FU14 combined year and mesh



Figure 3.8.4. Irish Sea East (FU14): Final discard ogive pooled for all years (2003-2014) and mesh sizes. L50=23.54 and L25=24.77, (IBPNeph 2015).

## Proportion males



Figure 3.8.5. Irish Sea East (FU14): Proportion of males in catch since 1999. Between 2010 and 2012 due to poor sampling levels estimates of sex ratio are not reliable. No sampling was carried out in 2020


Figure 3.8.6. Irish Sea East (FU14): Mean weight (g) combined by sex for total annual landings and discards. Values for 2010, 2011 and 2012 are not reliable due to poor sampling. NI sampling included from 2016. Average of 2017-2019 sampling used for 2020

CO3117 Grid - FU14


Figure 3.8.7. Irish Sea East (FU14): UWTV Survey stations for 2022. Highlighted stations 38 - 40 were not surveyed due to the presence of wind turbines.


Figure 3.8.8. Irish Sea East (FU14): Co-kriging approach. Interpolation result of VMS (cut off 3\%), survey density (20132015) data and mud distribution. A - model output; B - final polygon.



2016

406.1


2015


554

2017

544.4

2019



Figure 3.8.9. Irish Sea East (FU14): Burrow density estimates from the UWTV Survey (individuals / m².) Abundance estimates (millions) given at the bottom of each plot are adjusted with the cumulative absolute conversion factor (but does not contain the additional area for Wigtown Bay). Area of ground $=1032.75 \mathrm{Km}^{2}$ for $\mathbf{2 0 0 8} \mathbf{- 2 0 1 0}$ and $\mathbf{1 0 1 9 . 7 9 ~} \mathrm{Km}^{2}$ for 2011-2022.

Stock size


Figure 3.8.10. Irish Sea East (FU14): Abundance from the UWTV Survey. The shading indicates the $95 \%$ confidence interval. $B_{\text {trigger }}$ is set at $\mathbf{3 5 0}$ million (blue line).

## Fishing pressure



Figure 3.8.11. Irish Sea East (FU14): Harvest Rate (\% dead removed/UWTV abundance). The blue line indicates Fmsy proxy (11\%). Between 2010 and 2012, due to poor sampling levels, harvest rate estimates are not reliable.

# 16 Norway lobster (Nephrops norvgicus) in Division 7.a, Functional Unit 15 (Irish Sea, West) 

## Type of assessment

The assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the general process defined by WKNEPH (2009) described in the stock annex. The TV survey is due to be repeated in the summer of 2021 and forms the basis of advice for this stock in the autumn.

## ICES advice applicable to 2022

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2018-2020, catches in 2022 should be no more than 11785 tonnes.
To ensure that the stock in Functional Unit (FU) 15 is exploited sustainably, management should be implemented at the FU level.
ICES notes the existence of a management plan, developed and adopted by one of the relevant management authorities for Subarea 7. ICES considers this plan to be precautionary when implemented at the FU level."

### 16.1 General

## Stock description and management units

The Irish Sea West (FU15) is comprised of ICES rectangles 35E3-E5, 36E3-E5, 37E3-E5 and 38E4 within 7a. It is included in ICES Area 7 together with the Irish Sea East (FU14), Porcupine Bank (FU16), Aran Grounds (FU17) northwest Irish Coast (FU18), southeast and southwest Irish Coast (FU19), NW Labadie, Baltimore and Galley, and Jones and Cockburn (FU20-21) and the Smalls (FU22).
A TAC is in place for ICES Area 7 which does not correspond to the assessment units. As Nephrops are limited to muddy habitats the distribution of suitable sediment defines the species distribution and the stocks are therefore assessed as seven separate Functional Units. The TAC for Area 7 is shown in the tables below.

## Fishery description

The FU 15 Nephrops fishery first developed in the late 1950s. The environment in the Western Irish Sea is very suitable for Nephrops, with a large mud patch and a gyre that retains the larvae over the mud patch, thus ensuring good recruitment. The ground can be characterized as an area of very high densities of small Nephrops compared to other functional units. The UK (Northern Ireland) and Ireland are the main countries involved in the FU15 Nephrops fishery.

## The fishery in 2021

The Nephrops fishery in the Irish Sea west is economically one of the most important in ICES Area 7 and is mainly prosecuted by vessels from UK (Northern Ireland) and Ireland. Working Group landings from FU15 are presented in Table 16.1 and Figure 16.1. The total declared international

Nephrops landings reported from FU15 in 2021 was 6779 t , which was a decrease increase from 2019. The low levels of landings in 2019 ( 6115 t ) are considered to reflect behavioural changes due to the COVID-19 pandemic (Table 16.1). There has been a trend for Irish, since 2012, and more recently Northern Irish vessels to switch to multi (quad) rig trawls and in general a reduction of single-rig vessel effort. Since March 2012, it is mandatory for all Irish vessels to use specified species selective gears. Similar conditions have been introduced in October 2012 for the UK (Northern Ireland) vessels. The introduction of highly selective gears suggests a reduction in bycatch rates of non-target fish species of around $30 \%$. Quad-rig vessels are thought to increase Nephrops catch rates by around $30 \%$ whilst further reducing fish bycatch of $\sim 30 \%$ due to the lower headline height. In 2021 there was small increase in LPUE in 2021 for Northern Irish vessels whilst the LPUE for Irish vessels remained stable (Table 16.2).s

Further general information on the fishery can be found in the stock annex.

## Information from stakeholders

No information from stakeholders.

### 16.2 Data

Commercial size composition data for landings and discards were provided by Northern Ireland and Ireland. Other biological data used in the assessment were as listed in the stock annex compiled by the Benchmark meeting WKNEPH (2009).

## InterCatch

Data were available in InterCatch and used to derive assessment input data.

## Landings

Working Group landings from FU15 are presented in Table 16.1 and Figure 19.1. The total declared international Nephrops landings reported from FU15 in 2021 was 6779 t. Landings are derived primarily from Ireland and Northern Ireland. In 2021 there was significant impact of COVID-19 on the behaviour of fishers due to movement and social restrictions as well as market changes due to changing levels of demand.

## Effort

Effort by the UK fleet remained relatively stable since 2002 following a steady decline from the early 1990s. There was a further marginal reduction in effort and lpue time-series for Ireland (Table 16.3) compared to 2016, with effort at the lowest reported value in the series. In previous years these interannual fluctuations have been attributed to the high mobility and flexibility, in terms of fishing in other areas within the TAC area, whereas the Northern Irish effort is mostly concentrated on FU15. Fishing activity from the Irish fleet in FU15 increasingly concentrates on good fishing periods during the year, resulting in a larger and increasing lpue. The lpue and effort lpue series for Northern Ireland are updated to provide kW days $(\mathrm{kWd})$ and lpue as $\mathrm{kg} / \mathrm{kWd}$. A change to e-logbooks and recording of fishing hours after 2013 means that the recent data are not comparable with the historic series. Recent lpue and effort after 2013 has remained stable. The lpue for the Northern Irish and Irish fleets in 2021 were similar $2.71 \mathrm{~kg} / \mathrm{kWd}$ and $2.84 \mathrm{~kg} / \mathrm{kWd}$ but both increases since 2018 from $2.56 \mathrm{~kg} / \mathrm{kWd}$ and $2.7 \mathrm{~kg} / \mathrm{kWd}$.

## Sampling levels

Sampling of catches was impacted by COVID-19 in 2020 with cessation of sampling in the second quarter. Sampling resumed fully in 2021 and comparable levels to pre-2020 levels. Fisher selfsampling for Northern Irish vessels achieved 256 samples collected from the reference fleet, with $134,66,48$ and 8 samples in quarters $1-4$ respectively. The number of discard and catch samples collected from the Irish fleet was $0,10,22$ and 2 samples collected in quarters $1-4$ respectively. These rates correspond to one sample per 20.3 t landed by the Northern Irish fleet and one sample for every 44.4 t landed by the Irish fleet. Sampling levels due to changes in the schemes to adapted to COVID-19 impacts on working and social distancing requirements.

## Commercial length-frequency distributions

Length and sex compositions of Nephrops landed from the Irish Sea West are estimated from port sampling by Ireland and Northern Ireland. Sampling of Northern Ireland catches was not possible during 2003-2007, with the Irish length frequencies raised to the international catch for these years. Northern Ireland sampling resumed in 2008 and these data are combined with those from Ireland for that year.

This Northern Irish fisher self-sampling scheme uses a reference fleet of vessels selected vessels from the main Northern Irish ports. The reference vessels selection is designed to be representative of the entire fleet with systematic rota sampling. The mean sizes of Nephrops in the catches of both the Northern Ireland and Ireland fisheries have fluctuated for the last decade (Tables 19.4-19.5; Figure 16.1). There is little evidence to suggest a long-term trend in the mean size of males and females in the landings and catches which continues to fluctuate around the series mean (Figure 16.2).

## Sex ratio

The sex ratio by year is shown in Figure 16.3. This shows some fluctuations over time. In general, the sex ratio in landings and catches are biased toward males, with a geomean of $56.2 \%$ males in landings (1986-2020) and $52.4 \%$ in catches (1986-2020). There was little bias toward males in catches was observed in 2021 comprising $58 \%$ in landings and $52 \%$. Historically the stronger bias of males in landings relates to the average larger size of male Nephrops.

## Mean weights

Explorations of the mean weight in the catch samples by sex shows a strong seasonal pattern in the females (Figure 16.4). This corresponds with the emergence of mature females from the burrows to mate in summer. There is no evidence of a recent trend toward decreasing mean weights (Figure 16.5), however compared to the early part of the time-series mean weights have decreased. The mean weights in landings (2016-2021) and mean weights in discards (2016-2020) are used in the basis for calculating catch options (Section 19.4).

## Discards

Annual discard rates are estimated using unsorted catch and discards sampling. Unsorted catches and samples of retained catch are provided by vessels. The catch sample is partitioned into landings and discards using a discard selection ogive. This selection ogive can be derived per sample or as aggregation of samples within a quarter or year when sampling rates are low. Sampling effort is stratified weekly, but quarterly aggregations are used for quarterly length
frequencies and discard estimates. The length-weight regression parameters given in the stock annex are used to calculate sampled weights and appropriate raising factors. Discarding practice is highly variable, mainly driven by market demand, and was $26.5 \%$ of the catch by number in 2021 (Table 16.6). A discard survival rate of $10 \%$ is assumed for Nephrops from this FU (WKNEPH 2009).

## Surveys

## Abundance indices from UWTV surveys

Since 2003, Ireland and Northern Ireland have jointly carried out underwater television surveys of the main Nephrops grounds in the western Irish Sea. These surveys were based on a randomised fixed-grid design. The methods used during the surveys were similar to those employed for UWTV surveys of other Nephrops stocks and were as agreed by WKNEPHTV (ICES, 2007), WKNEPBID (ICES, 2008), SGNEPS (ICES, 2009; 2010; 2012), WKNEPH (ICES, 2009) and WGNEPS (ICES, 2013; 2014; 2015; 2016). From 2003 to 2011 year an average of 146 valid stations was covered by the two surveys combined, and the data were raised to a stock area of around $5290 \times 10-6 \mathrm{~km}^{2}$ as detailed in Table 16.7. The number of stations were significantly reduced in 2012 following a recommendation from SGNEPS 2012 that a CV (or relative standard error) of $<20 \%$ is an acceptable precision level for UWTV survey estimates of abundance. This allowed sampling intensity to be reduced and survey effort allocated to other areas and FUs in area 7. Details of the survey methodology are available in WGNEPS (ICES, 2016). Figure 16.6 shows the distribution of stations sampled in 2022. In 2022 the survey was completed on both the RV Corystes ( 82 stations) and Celtic Explorer (13 stations). In 2021 there was change to using High Definition 'still' image cameras to collect footage onboard the RV Corystes, this was used agin in 2022. This change provides significantly improved image quality. A similar change has also taken place in other functional units in ICES area 7. A trial in FU 16 showed no significant different in the burrow estimates derived from standard video imagery and high definition still imagery. In 202297 stations were completed, footage form 3 stations was not collected because of the presence of static fishing gear preventing the deployment of the camera. Figures 16.7-16.10 are contour plot of the krigged-density estimates for FU15 over the period 2003-2022. The resulting krigged burrow abundance estimate was 4498 million burrows. This was a similar result of that obtained in 2019 of 4775 million burrows. A violin plot of the burrow densities observed in the survey (2003-2022) is shown in Figure 16.11. The character of the burrow densities encountered has remained consistent over time; characterised by a relatively high occurrence of low density stations and a normal distribution densities around 0.74 burrow $/ \mathrm{m}^{2}$. Confidence in the survey estimates and design are assured through the maintained low coefficient of variation on the burrow estimates. This low coefficient of variation, despite the loss of three survey stations supports that the survey provides high quality information
The use of the UWTV surveys for the provision of Nephrops management advice was extensively reviewed by WKNEPH (ICES, 2009) and potential biases were highlighted including those due to edge effects; species burrow misidentification and burrow occupancy. A cumulative bias correction factor estimated for FU15 was 1.14 which means the TV survey is likely to overestimate Nephrops abundance by $14 \%$.

## Nephrops trawl surveys

In addition to UWTV surveys Northern Ireland have completed spring (April) and summer (August) Nephrops trawl surveys since 1994 and provide data on catch rates, size composition and biological data from fixed stations in the western Irish Sea as detailed in the Stock Annex (Stock Annex Figure 1). Survey cpue has remained stable over time. Mean carapace length-by-sex (from
the trawl survey) shows inter-annual variation fluctuating around mean with no apparent trend over time (Figure 16.12).

Due to reduced resources, the spring survey series was terminated in 2010 as part of a national rationalisation of the survey programme after considering benefits to management and stock assessment. Due to a major ship break-down, no data are available for the 2013 summer survey. The summer trawl survey catch rates correlate somewhat with UWTV survey abundance estimates (Figure 16.13), but showed a deviating trend, especially in 2010. The longer time-series of the trawl survey shows that catch rates in the last few years $(2005-2009,2011)$ are close to the mean of the series when UWTV burrow abundances were in the range of 5-6 billion burrows. The reduction in the 2010 trawl estimate, that showed a conflicting trend to the UWTV abundance, is most likely associated with the survey taking place in suboptimal tidal conditions. Usually the trawl survey coincides with slack tides, but this was not optimal in 2010 due to availability of the ship and synchronisation with the UWTV survey.

### 16.3 Assessment

## Comparison with previous assessments

The assessment approach used by WGCSE 2022 is consistent with that set out in the stock annex and WKNEPH (WKNEPH, 2009). Since the most recent three years of sampling data were available, three-year averages of mean weights in the landings and proportions retained in the fishery have been used. This is in line with the procedure used for other stocks in areas 6 and 7 by WGCSE.

## State of the stock

The stock size is estimated to show a decrease, but within the limits previously observed for the stock. The harvest ratio in 2021 (15.4\%) and remains below $\mathrm{Fmsy}^{\text {(18.1) (Figure 16.14). This stock }}$ has previously sustained landings at around 9000 t for many years. The stock increased until 2003, with a general decrease until 2014 and has increased since then. The most recent UWTV abundance estimate of 4498 million in 2022 follows a period (2016-2017) of above average size. The geometric mean of current series is 4908 million. Figure 16.14 is the stock summary plot for FU15. Recent harvest rates have fluctuated around FMSY, but is estimated as 15.4 in 2021, having decreased from 19.9 in 2015 (Table 16.6). The stock is estimated to be above MSYB ${ }_{\text {trigger }}$ (3000 million).

### 16.4 Catch option table

Catch option table inputs are presented in Table 16.6 and summarised below. A three year average (2019-2021) of mean weight in the landings and proportion of removals retained was used.
A stock abundance prediction for 2023 was made for FU15 using the approach agreed at the Benchmark Workshop (WKNEPH, 2009) and outlined in the stock annex made on the basis of the 2021 UWTV survey.

The basis for the catch options.

| Variable | Value | Notes |
| :--- | :---: | :--- |
| Stock abundance (2023) | 4498 | UWTV survey 2022; numbers of individuals in millions |
| Mean weight in projected <br> landings | 15.0 | Average 2019-2021; in grammes |
| Mean weight in projected <br> discards | 8.07 | Average 2019-2021; in grammes |
| Projected discard rate | 26.5 | Average 2019-2021; percentage by number of the total catch |
| Discard survival rate | 10.0 | Percentage by number of the discards |

### 16.5 Reference points

A decision-making framework for the choice of Fmsy proxy reference points is available in the introduction to the Nephrops ICES advice sheets. The current Fmsy proxy reference points for FU15 Nephrops was evaluated at WKMSYRef4. The MSY reference point for FU15 Nephrops is the Fmax for combined sexes. No precautionary reference points have been defined for Nephrops stocks. Whereas the FmSy proxy reference points were chosen with the intent that they should lead to a low probability of stock overfishing.

Previously the cpue data from the trawl surveys were scaled to the UWTV index to provide a $B_{\text {trigger }}$ approximation based on the mean of the five lowest survey catch rates in the time-series (Figure 16.8), this is still accepted as an appropriate Btrigger for FU15.

| Stock code | MSY Flower | F MSY | MSY Fupper with AR | MSY B $_{\text {trigger }}$ | MSY Fupper with no AR |
| :--- | :---: | :---: | :---: | :---: | :---: |
| nep-15 | 12.4 | 18.2 | 18.2 | $3000^{*}$ | 18.2 |

*Abundance in millions.

### 16.6 Management strategy

As yet there are no explicit management strategies for this stock.

### 16.7 Quality of assessment and forecast

Uncertainties in the survey, mean weight in the landings and discard rates are not taken into account in the deterministic catch option. There is some variability in these over time.

There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008; SGNEPS 2009). These have led to a revision in the historical time-series of survey abundance estimates for FU15, which was presented to last year's Working Group. Ultimately there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996).

Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that was more accurate but no more precise (WKNEPH 2009). The survey estimates themselves are very precisely estimated (CVs $2-5 \%$ ) given the homogeneous distribution of burrow density and the modelling of spatial structuring. The cumulative bias estimates for FU15 are largely based on expert opinion (see Stock Annex). The precision of these bias corrections cannot yet be characterised but is likely to be higher than that observed in the survey.

In the provision of catch options based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. These parameters are quite variable, in future years the uncertainty in these key parameters should be estimated.
The quality of landings data has improved since 2007 with the implementation of sales notes and buyers and sellers legislation. Prior to that there were concerns that landings were underreported. The harvest ratio may be under estimated prior to 2007.

### 16.8 Recommendations for next benchmark

WGCSE will keep the stock under review and recommend future benchmark as required.

### 16.9 Management considerations

The FU15 Nephrops fishery first developed in the late 1950s. Since then it has sustained landings of around 8500 t for more than 30 years. Fishing effort in the past has been very high but has declined somewhat in recent years. The environment in the Western Irish Sea is very suitable for Nephrops with a large mud patch and gyre, which retains the larvae over the mud patch thus ensuring good recruitment. The ground can be characterised as an area of very high densities of small Nephrops. All available information indicates that size structure of catches appears to have changed little since the fishery first began.

The Nephrops trawl fisheries take bycatches of other species, especially juvenile whiting, but also cod. Catches of these species should be reduced to as low as possible because of the poor status of these stocks. A conditional national licence has been introduced by Ireland since March 2012, making the use of grids or separator panels mandatory for all TR2 boats fishing in the Irish Sea. Around $55 \%$ of the Irish vessels use separator trawls and while $45 \%$ have opted to use Swedish grids to reduce bycatch. Additionally, there has been a trend for Irish vessels to switch to multi (quad) rig trawls. Provisional data suggest a $\sim 30 \%$ increase in Nephrops catch rates and a reduction in fish bycatch of $\sim 30 \%$ due to the lower headline height.

Since October 2012, all TR2 vessels in the UK (Northern Ireland) fleet are required to use a highly selective fishing gear. In the Irish Sea these currently include Seltra 300 mm box trawl, 270 mm diamond mesh panel Seltra box trawl and 300 mm square mesh panel. All these gears are being developed with the aim of achieving exemption from the cod recovery plan under Article 11 (less than $1.5 \%$ cod catch). Enforcement is through the issue cod recovery zone fishing authorisations, where no authorisation is given to a vessel that is not using a highly selective gear.
ICES has repeatedly advised that management should be at a smaller scale than the ICES Subarea 7. Management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort are at the same scale as the resource.
A number of cod recovery measures have been introduced since 2000 to promote recovery of Irish Sea cod stocks. These include a closure of the western Irish Sea cod spawning grounds from mid-February to end of April since 2000, with a later extension to the eastern Irish Sea closure. Despite a partial derogation for Nephrops vessels during the closed period the distribution of effort on Nephrops has been affected by this management plan. There have also been
decommissioning schemes to reduce fishing effort. During 2016-2020 the EU landing obligation was applied to all catches of Norway lobster fisheries in ICES Subarea 7 with exemptions for high survival. From 2020, this stock is still under a landing obligation and there are still exemptions in place. Observations from the 2016-2020 fishery indicate that discarding above the minimum conservation reference size (MCRS) continues and has not changed markedly (Figure 3). ICES is providing advice for 2022 assuming average discard rates as observed over the last three years. This is considered to be the most realistic assumption.

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Table 16.1. Irish Sea West (FU15): Landings (tonnes) by country, 2000-2018.

| Year | Ireland | Isle of Man | UK | Other countries | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 3433 | 0 | 4937 | 0 | 8370 |
| 2001 | 2689 | 3 | 4749 | 0 | 7441 |
| 2002 | 2291 | 1 | 4501 | 0 | 6793 |
| 2003 | 2709 | 4 | 4352 | 0 | 7065 |
| 2004 | 2786 | 13 | 4470 | 1 | 7270 |
| 2005 | 2133 | 0 | 4420 | 0 | 6554 |
| 2006 | 2051 | 1 | 5508 | 1 | 7561 |
| 2007 | 2767 | 0 | 5724 | 0 | 8491 |
| 2008 | 3132 | 50 | 7323 | 2 | 10508 |
| 2009 | 2343 | 1 | 6855 | 0 | 9198 |
| 2010 | 2578 | 0 | 6384 | 0 | 8963 |
| 2011 | 3575 | 2 | 6584 | 0 | 10162 |
| 2012 | 3794 | 3 | 6732 | 0.2 | 10529 |
| 2013 | 2465 | 31 | 6175 | 0.2 | 8672 |
| 2014 | 2938 | 0** | 5676 | 0.0 | 8613 |
| 2015 | 2199 | 0** | 6433 | 0.3 | 8632 |
| 2016 | 1609 | 0** | 5715 | 3 | 7327 |
| 2017 | 1253 | 0** | 4896 | 0 | 6150 |
| 2018 | 1387 |  | 4369 | 0 | 5756 |
| 2019* | 1859 |  | 5731 | 0 | 7590 |
| 2020 | 1555 |  | 4560 |  | 6115 |
| 2021 | 1512 |  | 5267 |  | 6779 |

* provisional. **included in UK landings.

Table 16.2. Irish Sea West (FU15): Catches and landings (tonnes), effort ('000 hours trawling), cpue and lpue (kg/hour trawling) Republic of Ireland Nephrops Directed Trawlers 2000-2019.

| Year | Landings ( Kg ) | Effort (Hours) | Effort (days) | Effort (kwdays) | Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 1706969 | 44459 | 3516 | 835977 | 2.041885 |
| 1996 | 1406140 | 31409 | 2326 | 607785 | 2.313549 |
| 1997 | 2801501 | 60502 | 4518 | 1124379 | 2.491599 |
| 1998 | 2696979 | 52277 | 4051 | 1053491 | 2.560039 |
| 1999 | 4031508 | 73786 | 5260 | 1367903 | 2.947217 |
| 2000 | 3227565 | 61936 | 4396 | 1199896 | 2.68987 |
| 2001 | 2428587 | 51111 | 3435 | 939387 | 2.585289 |
| 2002 | 2015965 | 46072 | 2900 | 873563 | 2.307749 |
| 2003 | 1620391 | 47704 | 3120 | 878568 | 1.844355 |
| 2004 | 2586760 | 52673 | 3500 | 1033073 | 2.503946 |
| 2005 | 2111185 | 50825 | 3414 | 1003901 | 2.102981 |
| 2006 | 2031881 | 53461 | 3535 | 1084251 | 1.873995 |
| 2007 | 2728841 | 52550 | 3575 | 1056291 | 2.583419 |
| 2008 | 3165781 | 49218 | 3401 | 1027919 | 3.079796 |
| 2009 | 2333433 | 34651 | 2368 | 706178 | 3.304312 |
| 2010 | 2505061 | 36504 | 2546 | 739345 | 3.388218 |
| 2011 | 3554343 | 47640 | 3229 | 921298 | 3.857972 |
| 2012 | 3725318 | 49313 | 3560 | 966006 | 3.856413 |
| 2013 | 2269336 | 33818 | 2571 | 682793 | 3.323608 |
| 2014 | 2449612 | 40371 | 3007 | 852740 | 2.872635 |
| 2015 | 2119880 | 35898 | 2733 | 756719 | 2.80141 |
| 2016 | 1529418 | 28249 | 2301 | 556452 | 2.748516 |
| 2017 | 1120690 | 22516 | 1749 | 410628 | 2.729208 |
| 2018 | 1363911 | 27084 | 1919 | 535002 | 2.549353 |
| 2019 | 1803134 | 33981 | 2304 | 700132 | 2.57542 |
| 2020 | 1517909 | 25717 | 2250 | 570314 | 2.661534 |
| 2021 | 1517909 | 25717 | 2250 | 570314 | 2.661534 |

Table 16.3. Irish Sea West (FU15): Landings (tonnes), effort ('000 hours trawling), lpue (kg/hour trawling), effort ('000 kW days) and Ipue (kg/kWd) of Northern Ireland Nephrops trawlers, 2000-2019.

| Year | Landings | Effort ('000 hours) | Ipue ('000 hrs) | kW days ('000) | Ipue kWd |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 4758 | 168.7 | 28.2 |  |  |
| 2001 | 4587 | 163.7 | 28.0 |  |  |
| 2002 | 4495 | 130.8 | 34.4 |  |  |
| 2003 | 4146 | 136.1 | 29.0 |  |  |
| 2004 | 4273 | 144.3 | 29.6 |  |  |
| 2005 | 4235 | 138.4 | 30.6 |  |  |
| 2006 | 5356 | 144.1 | 37.2 |  |  |
| 2007 | 5512 | 126.9 | 43.4 |  |  |
| 2008 | 7056 | 141.4 | 49.9 |  |  |
| 2009 | 6487 | 134.7 | 48.2 |  |  |
| 2010 | 5888 | 141.1 | 41.7 |  |  |
| 2011 | 5952 | 132.7 | 44.9 |  |  |
| 2012 | 5865 | 137.8 | 42.6 |  |  |
| 2013 | 5605 | 135.7 | 41.3 | 2151.9 | 2.60 |
| 2014 | 5190 | 114.6 | 45.3 | 2111.2 | 2.46 |
| 2015 | 6396 |  |  | 1962.6 | 3.26 |
| 2016 | 5638 |  |  | 2107.3 | 2.68 |
| 2017 | 4789 |  |  | 1904.3 | 2.51 |
| 2018 | 4293 |  |  | 2079.3 | 2.06 |
| 2019 | 5539 |  |  | 2166.5 | 2.56 |
| 2020 | 4550 |  |  | 1852.0 | 2.46 |
| 2021 | 5201 |  |  |  |  |

[^9]Table 16.4. Irish Sea West (FU15): Mean sizes (mm CL) of male and female Nephrops in Northern Ireland catches, landings and discards, 2000-2018.

| Year | Catches |  | Landings |  | Discards |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Males | Females | Males | Females | Males | Females |
| 2000 | 27.7 | 24.5 | 29.4 | 26.3 | 22.5 | 22.6 |
| 2001 | 25.7 | 23.6 | 26.1 | 24.4 | 21.7 | 21.2 |
| 2002 | 26.7 | 24.1 | na | 26.7 | 24.9 | 21.8 |
| 2003 | na | na | na | na | na | na |

na $=$ not available.

Table 16.5. Irish Sea West (FU15): Mean sizes (mm CL) of male and female Nephrops in Republic of Ireland catches, landings and discards, 2000-2018.

| Year | Catches |  | Landings |  | Discards |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Males | Females | Males | Females | Males | Females |
| 2000 | 29.1 | 27.1 | 32.2 | 29.7 | 24.3 | 24.0 |
| 2001 | 26.7 | 24.8 | 28.6 | 27.0 | 23.0 | 22.2 |
| 2002 | 28.9 | 25.4 | 30.2 | 27.8 | 24.6 | 23.6 |
| 2003 | 27.7 | 24.9 | 29.7 | 26.9 | 24.0 | 23.1 |
| 2004 | 28.1 | 26.1 | 29.7 | 27.8 | 23.9 | 23.7 |
| 2005 | 28.5 | 26.8 | 30.1 | 29.1 | 23.9 | 23.2 |
| 2006 | 27.7 | 25.5 | 29.5 | 27.1 | 23.8 | 23.1 |
| 2007 | 27.7 | 25.4 | 29.8 | 27.9 | 24.0 | 23.3 |
| 2008 | 27.4 | 24.6 | 28.9 | 26.6 | 22.0 | 21.4 |
| 2009 | 28.5 | 26.3 | 30.5 | 29.2 | 24.3 | 23.4 |
| 2010 | 28.0 | 25.9 | 29.6 | 27.6 | 23.8 | 23.3 |
| 2011 | 27.0 | 25.7 | 28.8 | 27.3 | 23.7 | 23.5 |
| 2012 | 26.8 | 25.6 | 28.3 | 27.0 | 23.2 | 23.0 |
| 2013 | 26.3 | 25.1 | 27.4 | 26.5 | 23.1 | 22.6 |
| 2014 | 27.7 | 24.9 | 29.2 | 26.3 | 23.6 | 23.3 |
| 2015 | 27.7 | 25.7 | 29.5 | 27.4 | 24.4 | 24.0 |
| 2016 | 26.0 | 25.0 | 27.3 | 26.4 | 23.5 | 23.3 |
| 2017 | 27.2 | 25.0 | 28.1 | 26.2 | 23.4 | 22.6 |
| 2018 | 27.4 | 24.9 | 29.8 | 22.8 | 24.6 | 22.8 |
| 2019 | 27.9 | 25.0 | 29.5 | 27.0 | 22.8 | 22.3 |
| 2020 | 28.0 | 26.3 | 29.7 | 27.9 | 24.1 | 24.1 |
| 2021 | 27.9 | 25.8 | 29.6 | 28.4 | 23.2 | 23.9 |

Table 16.6. Irish Sea West (FU15): Proportion discarded by weight and number from FU15. (Note a 10\% survivorship of discards is assumed in HR and forecast calculations).

|  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Millions |  | millions |  |  | \% | tonnes |  | \% |  | grammes |  |
| 2003 | 5485 | 0.027 | 404 | 291 | 666 | 12.1 | 7065 | 2659 | 41.9 | 39.3 | 17.5 | 9.14 |
| 2004 | 5547 | 0.03 | 416 | 218 | 612 | 11.0 | 7270 | 1993 | 34.4 | 32.0 | 17.5 | 9.14 |
| 2005 | 5673 | 0.044 | 346 | 157 | 488 | 8.6 | 6554 | 1412 | 31.2 | 29.1 | 18.9 | 8.99 |
| 2006 | 5402 | 0.041 | 467 | 261 | 701 | 13.0 | 7561 | 2285 | 35.9 | 33.4 | 16.2 | 8.75 |
| 2007 | 5150 | 0.034 | 511 | 375 | 848 | 16.5 | 8491 | 3246 | 42.3 | 39.7 | 16.6 | 8.66 |
| 2008 | 4288 | 0.025 | 755 | 191 | 927 | 21.6 | 10508 | 1421 | 20.2 | 18.6 | 13.9 | 7.44 |
| 2009 | 4623 | 0.026 | 567 | 335 | 868 | 18.8 | 9198 | 2934 | 37.1 | 34.7 | 16.2 | 8.76 |
| 2010 | 4990 | 0.031 | 572 | 180 | 733 | 14.7 | 8963 | 1539 | 23.9 | 22.0 | 15.7 | 8.55 |
| 2011 | 4871 | 0.023 | 644 | 332 | 943 | 19.4 | 10162 | 2683 | 34.0 | 31.7 | 15.8 | 8.08 |
| 2012 | 5062 | 0.029 | 771 | 258 | 1003 | 19.8 | 10529 | 1871 | 25.1 | 23.1 | 13.7 | 7.25 |
| 2013 | 4310 | 0.027 | 662 | 229 | 867 | 20.1 | 8672 | 1590 | 25.7 | 23.6 | 13.1 | 6.94 |
| 2014 | 4593 | 0.025 | 641 | 198 | 819 | 17.8 | 8613 | 1418 | 23.6 | 21.7 | 13.4 | 7.16 |
| 2015 | 4373 | 0.029 | 620 | 280 | 872 | 19.9 | 8643 | 2228 | 31.1 | 28.9 | 13.9 | 7.96 |


|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \stackrel{i}{\pi} \\ & \stackrel{\text { ® }}{\sim} \end{aligned}$ |  | Millions |  | millions |  |  | \% | tonnes |  | \% |  | grammes |  |
| 2016 |  | 5076 | 0.03 | 562 | 245 | 783 | 15.4 | 7327 | 1939 | 30.4 | 28.2 | 13.0 | 7.91 |
| 2017 |  | 5312 | 0.03 | 426 | 152 | 563 | 10.6 | 6150 | 1222 | 26.3 | 24.3 | 14.4 | 8.04 |
| 2018 |  | 4932 | 0.03 | 360 | 145 | 491 | 10.0 | 5756 | 1231 | 28.7 | 26.6 | 16.1 | 7.43 |
| 2019 |  | 4404 | 0.03 | 536 | 154 | 675 | 15.3 | 7590 | 1159 | 22.3 | 20.5 | 14.2 | 7.54 |
| 2020 |  | 4775 | 0.03 | 371 | 153 | 509 | 10.7 | 6115 | 1294 | 29.2 | 27.1 | 15.5 | 8.64 |
| 2021 | 4733 | 0.04 |  | 440 | 172 | 595 | 12.4 | 6779 | 1379 | 28.1 | 26.0 | 15.4 | 8.02 |
| 2022 | 4498 | 0.03 |  |  |  |  |  |  |  |  |  |  |  |

Table 16.7. Irish Sea West (FU15): Results from NI/ROI collaborative UWTV surveys of Nephrops grounds in 2003-2020.

| Ground | Year | Number of stations | Mean Density ( $\mathrm{No} . / \mathrm{M}^{2}$ ) | Domain Area (km²) | Estimate <br> (billions) | CV on Burrow estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2003 | 160 | 0.99 | 5295 | 5.5 | 3\% |
|  | 2004 | 147 | 1.00 | 5310 | 5.5 | 3\% |
|  | 2005 | 141 | 1.02 | 5281 | 5.7 | 4\% |
|  | 2006 | 138 | 0.97 | 5194 | 5.4 | 4\% |
|  | 2007 | 148 | 0.93 | 5285 | 5.1 | 3\% |
|  | 2008 | 141 | 0.77 | 5287 | 4.3 | 3\% |
|  | 2009 | 142 | 0.83 | 5267 | 4.6 | 3\% |
|  | 2010 | 149 | 0.90 | 5307 | 5.0 | 3\% |
|  | 2011 | 156 | 0.88 | 5289 | 4.9 | 2\% |
|  | 2012 | 99 | 0.91 | 5291 | 5.1 | 3\% |
|  | 2013 | 80 | 0.78 | 5278 | 4.3 | 3\% |
|  | 2014 | 99 | 0.83 | 5272 | 4.6 | 3\% |
|  | 2015 | 100 | 0.79 | 5279 | 4.4 | 3\% |
|  | 2016 | 100 | 0.84 | 5260 | 5.1 | 3\% |


| Ground | Year | Number of stations | Mean Density ( $\mathrm{No} . / \mathrm{M}^{2}$ ) | Domain Area (km²) | Estimate (billions) | CV on Burrow estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2017 | 101 | 0.90 | 5304 | 5.3 | 3\% |
|  | 2018 | 100 | 0.85 | 5791 | 4.9 | 3\% |
|  | 2019 | 100 | 0.76 | 5370 | 4.4 | 3\% |
|  | 2020 | 99 | 0.82 | 5791 | 4.8 | $3 \%$ |
|  | 2021 | 95 | 0.78 | 5790 | 4.7 | 4\% |




Figure 16.2. Irish Sea West (FU15): Length distributions in the landings (solid) and catches (dotted) 1986-2022.


Figure 16.3 Nephrops in FU15 (Irish Sea West). Sex ratio (percentage of males) of landings and discards (1986-2020).


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Figure 16.4 Nephrops in FU15 (Irish Sea West).Mean weight in catch samples by sex with GAM loess smoother ( $\mathbf{k}=\mathbf{2 0}$ ).


Figure 16.5 Nephrops in FU15 (Irish Sea West). Mean weight in landings and discards.

## 2022 FU15 UWTV Density



Figure 16.6. Irish Sea West (FU15): 2022 UWTV survey stations, symbol size reflects the burrow density.


Figure 16.7. Irish Sea West (FU15): Contour plots of the krigged density estimates for the Irish Sea from 2003-2008.


Figure 16.8. Irish Sea West (FU15): Contour plots of the krigged density estimates for the Irish Sea from 2009-2014


Figure 16.9. Irish Sea West (FU15): Contour plots of the krigged density estimates for the Irish Sea from 2009-2020.


Figure 16.10. Irish Sea West (FU15): Contour plots of the krigged density estimates for the Irish Sea for 2021 and 2022


20032004200520062007200820092010201120122013201420152016201720182019202020212022

Figure 16.11. Irish Sea West (FU15): Box and kite plot of burrow density observed during UWTV survey 2003-2022.


Figure $\mathbf{1 6 . 1 2}$ Irish Sea West (FU15): Nephrops catches (kg per nm) from NI trawl surveys. No data available in 2013 due to ship breakdown.


Figure 16.13. Irish Sea West (FU15): Revised UWTV index and scaled trawl survey. Cpue along with $B_{\text {trigger }}$ based upon mean of five lowest trawl survey values. Abundance figures have not been bias corrected.


Figure 16.14 Norway lobster in Division 7.a, Functional Unit 15. Summary of the stock assessment. Catches (discard data are only available from 1986), harvest rate (sum of landings and dead discards in numbers, divided by total abundance), survey abundance (Underwater TV, billions; SSB proxy; 95\% confidence intervals). Harvest rates between 2003 and 2006 may be underestimated because of underreporting of landings. Orange lines represent MSY Btrigger and the Fmsy harvest rate.

# 17 Norway lobster (Nephrops norvegicus) in divisions 7.b-c and 7.j-Km Functional Unit 16 (west and southwest of Ireland) 

## Type of assessment in 2022

Available data on the fishery for 2021 and other stock indicators have been updated here according to the stock annex (Nephrops FU16). The assessment and catch options follow the agreed procedures set out in the stock annex; however, mean weight calculations for 2020 and 2021 deviated from the stock annex and are detailed in 17.4 Data section below.

## ICES advice applicable to 2021

"ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, and assuming zero discards, catches in 2021 that correspond to the F ranges in the MAP are between 2653 tonnes and 3290 tonnes. The entire range is considered precautionary when applying the ICES advice rule.
To ensure that the stock in Functional Unit (FU) 16 is exploited sustainably, management should be implemented at the functional unit level."

## ICES advice applicable to 2022

"ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, and assuming zero discards, catches in 2022 that correspond to the F ranges in the MAP are between 2261 and 2804 tonnes. The entire range is considered precautionary when applying ICES advice rule.
To ensure that the stock in FU 16 is exploited sustainably, management should be continued at the FU level."

### 17.1 General

## Stock description and management units

The TAC area is Subarea 7, since 2011 an 'of which' clause was introduced specifically for the Porcupine Bank (FU16) see Table 20.1. The Functional Unit for assessment includes some parts of the following ICES divisions $7 . \mathrm{b}, \mathrm{c}, \mathrm{j}$, and k . The exact stock area is shown on the map below and includes the following ICES Statistical rectangles: 31-35 D5-D6; 32-35 D7-D8.


The FU16 outlined by the red line. The closed area from 1 May-31 July since 2010 (reduced to only May since 2013) is shown with a green line. Irish Nephrops directed fishing effort between 2006-2009 derived from integrated VMS and logbook information is shown as a heat map.

## Management applicable to 2021 and 2022

## TAC in 2021

Council Regulation (EU) 2021/92 of 28 January 2021 fixing for 2021 the fishing opportunities for certain fish stocks and groups of fish stocks, applicable in Union waters and, for Union fishing vessels, in certain non-Union waters.


Functional Unit 16 of ICES Subarea 7 (NEP/*07U16):

| Spain | 992 |
| :--- | ---: |
| France | 621 |
| Ireland | 1194 |
| Union | 2807 |
| United Kingdom | 483 |
| TAC | 3290 |

TAC in 2022
Council Regulation (EU) 2022515 of 31 March 2022 amending Regulation (EU) 2022109 fixing for 2022 the fishing opportunities for certain fish stocks and groups of fish stocks applicable in Union waters and for Union fishing vessels in certain non-Union waters.

| Species: | Norway lobster <br> Nephrops norvegicus |  |  | Zone: | $\begin{aligned} & 7 \\ & (\mathrm{NEP} / 07 .) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Spain | $924{ }^{\text {(1) }}$ |  |  | Analytical TAC |  |
| France | 3746 |  |  |  |  |
| Ireland | 5682 (1) |  |  |  |  |
| Union | 10352 (1) |  |  |  |  |
| United Kingdom | 6686 |  |  |  |  |
| TAC |  | 17038 | (1) |  |  |
| (1) | Special condition: within the limits of these q taken in the following zone: <br> Functional unit 16 of ICES subarea 7 (NEP/*07U16) |  |  | otas, n | han the qu |
|  |  |  |  |  |  |
|  | Spain | 846 |  |  |  |
|  | France | 530 |  |  |  |
|  | Ireland | 1016 |  |  |  |
|  | Union | 2392 |  |  |  |
|  | United Kingdom | 412 |  |  |  |

### 17.2 Closed area restrictions

A seasonal closed area has been in place for three months May 1-31 July between 2010-2012 (shown in the map above and coordinates below). The period of the EU regulatory closure was reduced to only one month between 2013 and 2019 (Council Regulation 2019/124, Article 13).

The following TCMs are in place for Nephrops in 7 (excluding 7.a) after EC 850/98 in operation since 2000. Minimum Landing Sizes (MLS); total length $>85 \mathrm{~mm}$, carapace length $>25 \mathrm{~mm}$, tail length $>46 \mathrm{~mm}$. Although it is legal to land smaller prawns from this fishery, marketing restrictions imposed by producer organizations in France mean smaller Nephrops ( $<35 \mathrm{~mm}$ CL or 115 mm whole length) are not retained in this fishery.

The mesh size restrictions apply to towed gears in 7.b-k targeting Nephrops and are given in Section 7.1. Vessels mainly used $80-99 \mathrm{~mm}$ mesh to target Nephrops on the Porcupine Bank.

The landing obligation applied since 2016 for certain vessels that matched the criteria set out in the discard plans: https://ec.europa.eu/fisheries/cfp/fishing_rules/discards_en

### 17.3 Fishery in 2021

WGCSE reviewed effort trends for Irish vessels that accounted for $65 \%$ of the total landings in 2021. The Irish fishery in 2021 took place up to April, after which the fishery was closed, but was reopened from October to December. In 2017 the industry reported very good catches of Nephrops but commented that the mean size declined significantly; however, mean sizes increased in 2018 and 2019, decreased again in 2020 and increased in 2021 (Figure 20.4).

## Effect of regulations

Prior to 2011 TACs and quotas were applied to the whole Subarea 7, so the FU16 fishery was not restricted by quotas. Since 2011, the "of which clause" was implemented in the TAC regulation specifically for the Porcupine Bank. Quotas have been very restrictive for Irish vessels and this has led to various changes in fishing patterns. Vessels have tried to optimise the economic value of the catch by targeting areas and periods with relatively smaller ${ }^{1}$ volumes of larger higher value Nephrops. The FU16 specific quota has also increased area misreporting in the past and the risk of discarding. An unallocated component related to area misreporting was included in the assessment from 2011 to 2017. Since 2018, following the implementation of new legislation limiting fishing trips to single functional units, misreporting was not included in the assessment.

## Information from stakeholders

The provision of grade information by individual fishers and coops remains a highly important assessment input. However, in 2020 and 2021 graded information was not used in the assessment.

| Year | \% of Irish landings where grade data were provided |
| :--- | :--- | :--- |
| 2011 | $60 \%$ |
| 2012 | $45 \%$ |
| 2013 | $57 \%$ |
| 2014 | $33 \%$ |
| 2015 | $44 \%$ |
| 2016 | $49 \%$ |
| 2017 | $31 \%$ |
| 2018 | $31 \%$ |
| 2019 | $50 \%$ |

The industry collaborated with the development of an Irish Fisheries Science Research Partnership survey in 2010 (Stokes and Lordan, 2011).

The Irish industry considers that the stock has increased significantly and no longer requires the Functional Unit "of which" clause.

[^10]
### 17.4 Data

## InterCatch

Data were available in InterCatch and used on a trial basis.

## Landings

Total international landings increased by $30 \%$ in 2021 to 2476 t (Figure 20.1 and Table 20.2). From 2011 to 2017 total landings for FU16 had included "unallocated landings" from other FU due to misreporting. Since 2018 no reallocation has been applied as there was no information concerning misreporting.

## Sampling levels

Sampling levels, data aggregating and raising procedures were reviewed by WKNEPH 2013, and are documented in the stock annex. Recent sampling rate is provided in Table 20.3.

Since 2010 landings length distributions have been reconstructed using the methods outlined in the stock annex. This involves using samples of the grade length structure from Irish sampling and estimates of the volume of each commercial size grade provided by the fishing industry. This was used to reconstruct Irish LFDs; landings by other fleets, which accounted for $\sim 23 \%$ of the total landings from 2010 to 2019, were unsampled.

In 2020 and 2021, due to the low sampling levels of graded landings caused by COVID-19 restrictions, efforts were made to adapt the sampling programme. Unsorted catch samples were collected from five and 12 Nephrops fishing trips, respectively in 2020 and 2021.

## Commercial length-frequency distributions

The time-series of raised international length-frequency distributions of the sampled landings by sex are given in Figure 20.2. This also shows significant shift towards larger individuals in the landings between 2002-2009 when few individuals at smaller sizes were observed. The length distribution in 2019 was similar to 2018. The mean lengths by sex and year are presented in Table 20.4. These figures and tables are not updated for 2020 and 2021.

## Sex ratio

Previous Nephrops working groups have highlighted stability in sex ratio as an important indicator for Nephrops stocks. The landings and fishery-independent survey catches show a dramatic switch in the sex ratio for this stock with larger proportions of females in the catches of 2008 and 2009 (Figure 20.3). Both the commercial and survey data indicate that sex ratio switched back to a more usual situation since 2010 with males accounting for larger proportions of the catch/landings.

Nephrops moult once a year shortly after hatching of eggs in April or May. There is a 24 -hour period after moulting when the male Nephrops can mate with the female (Farmer, 1974). If there are insufficient males in the population to mate with the recently moulted females, this can result in a change in female behaviour whereby unmated females concentrate on feeding and growth instead of reproduction. This so called "sperm limitation" hypothesis could explain the sex ratio changes observed in the Porcupine Nephrops. WKNEPH 2013 examined the available scientific data on proportions of females mated observed on the Spanish survey. These results showed
high proportions of unmated females and a high L50 for mated females in catches in 2009. Simulations were also carried out to investigate the densities at which sperm limitation may become an issue given plausible ranges of stock density, sex ratios, search radii. The conclusion was that at the densities recently observed on the Porcupine Bank that sperm limitation was a real possibility.

## Mean weight explorations

The mean weights in the landings are shown for the full time-series in Figure 20.4 and Table 20.5. In 2020 and 2021, due to COVID-19 restrictions, mean weight calculations deviated from the stock annex and were estimated using the average mean weights of catch samples from five Nephrops fishing trips.

## Discards

There are few historical estimates of discards for this stock. Irish sampling up to 2016 observed very minimal discarding (mainly limited to small and damaged individuals $<5 \%$ by number). Four Irish trips were sampled in 2016. Discards were not recorded on one of these trips. However, on the other three trips, discards were estimated to be around $8 \%, 9 \%$ and $15 \%$ by number $(3 \%, 3 \%$ and $6 \%$ by weight). In 2017 there were two trips where discards were recorded, $17 \%$ and $43 \%$ by number. In 2018 discards were observed on one of the two trips ( $74 \%$ by number) no discards were observed on the other trip. In 2019, discards were observed in two of the four trips ( $13 \%$ and $29 \%$ by number). In 2020 discards were observed in two of the five trips. In 2021 discards were observed in four of the 12 trips. The discarding observed on these trips is likely not reflective of the overall discard pattern as the skippers advised the scientist on board that they had increased their discards to remain within quota during the observed trip. This means that the current discard pattern is unknown, but can be no longer considered negligible.

A detailed examination of discard estimates was provided in Spain in 2014. No estimate was provided in InterCatch by Spain since 2015.

## Abundance indices from UWTV surveys

Operational details of the 2022 UWTV survey are available (Aristegui et al., 2022). These surveys use the standard UWTV methodology and conforms to WGNEPS best practice and guidelines, documented in Dobby H., et al., 2021. WKNEPH 2013 recommended that these surveys could be used for assessment and provision of catch options. The results are given in Table 20.6. Further detail of the survey is provided in the annex and annual survey reports are available at http://oar.marine.ie/handle/10793/59.

## Trawl surveys

The longest time-series of fishery-independent source of data is from the Spanish Porcupine trawl survey 2001-2021 (SpPGFS-WIBTS-Q4). This survey is carried out in September when Nephrops catchability is quite low, particularly of adults. Further information on this survey is provided in the IBTS report (ICES, 2015) and in previous IBTS reports. Data from 2021 survey for this report was taken from Velasco et al. (in draft).
Distribution of Nephrops catches and biomass in Porcupine surveys between 2001 and 2021 is shown in Figure 20.5. There was a year effect in 2008 when unusual gear parameters were observed. Catch rates in 2011 may also have been reduced due to exceptionally poor weather and gear performance issues. The stratified abundance estimate and biomass increased significantly from 2015 to 2018, decreased in 2019 and 2020, and increased again in 2021 (Figure 20.6).

The size structure of the catches in the survey shows two things: a lower mean size than in the commercial fleets and an increasing trend in mean size for both sexes up to 2008. In 2009, there is large reduction of mean size in both sexes due to a recruiting year class with a modal length at around 27 mm (possibly the 2006 year class). Tracking of cohorts was carried out at WKNEPH 2013 but the results are inconclusive (ICES, 2013). The survey showed increased recruitment between 2013 and 2019 with significantly increased catch rates of individuals $<24 \mathrm{~mm}$ (Figure 20.7). This has also led to increase catch rates of juveniles and adult Nephrops since 2016.

An Irish Fisheries Science Research Partnership (IFSRP) survey was developed in collaboration with the Irish fishing industry to obtain data from the closed area in 2010-2012. Details of the design and methodology are presented in Stokes and Lordan (2011). The survey uses both commercial gear (Comm) and a baca trawl similar to the SpPGFS-WIBTS-Q4. WKNEPH concluded that the IFSRP trawl survey is too short (with changes in coverage, gears and vessels) to draw an inference about cpue changes reflecting changing stock abundance (ICES, 2013). The surveys carried out between 2010-2012 provided very useful data on population structure across the ground as well as data on grade structure and maturity-at-length.

## Commercial cpue

In the past the Nephrops fishery on the Porcupine Bank was both seasonal and opportunistic with increased targeting during periods of high Nephrops emergence and good weather. Freezing of catches at sea has become increasingly prevalent since 2006, and the fishery now operates throughout the year, mainly targeting larger more valuable Nephrops in lower volumes. Fishing effort has fluctuated considerably in the recent past in response to availability of Nephrops.
Effort and lpue/cpue data are generally not standardized, and hence do not take into account vessel capacity, efficiency, seasonality or other factors that may bias perception of lpue/cpue and abundance trends over the longer term. WKNEPH concluded that effort and lpue series should be maintain in the WGCSE report for information purposes (ICES, 2013). WGCSE 2016 recommended presenting the effort in KWDays and lpue in tonnes/ KWDays. Any inferences about changes in stock abundance from these data, should take account of the quality and bias concerns raised above.

These data are presented by country in Table 20.7.

### 17.5 Stock assessment

## Comparison with previous assessments

This assessment is based on UWTV approach outlined in WKNEPH 2013 and using parameter in the stock annex (ICES, 2013). This year's assessment has been updated based on the results of the August 2022 UWTV survey (Aristegui et al. 2022).

## State of the stock

The UWTV results are shown in Table 20.6. In 2017 the harvest rate was above FMSY for the first time. However, since 2018 the harvest rate has been below FMSY again, due to relatively high abundance estimates on UWTV surveys since 2018, and to the increase in mean weight in the landings, which resulted in a decrease in the landed numbers. Total abundance increased in 2022, and it is the highest value in the time-series.

## Catch options table

The inputs to the catch options are given below. At this point, it is not possible to estimate the numbers and mean weights of discards in the fishery, although there are indications that discards have increased since 2016.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Stock abundance (2023) | 1363 | UWTV survey 2022; numbers of individuals in millions |
| Mean weight in projected landings | 44.8 | Average 2019-2021; in grammes |
| Mean weight in projected discards | - | Unknown |
| Projected discard rate | - | Unknown |
| Discards survival rate | - | Not applicable |

### 17.6 Reference points

New reference points were evaluated by WKMSYREF4 (ICES, 2016a) and advised by ICES (2016b). The Fmsy for this stock was increased from $5.0 \%$ to $6.2 \%$. The Fmsy for this stock is based on $\mathrm{F}_{0.1}$ for both sexes combined given the low density of Nephrops on the Porcupine Bank.

| Stock code | MSY Flower* | F MSY $^{*}$ | MSY F upper $^{*}$ with AR | MSY B trigger | MSY Fupper $^{*}$ with no AR |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| nep-16 | $5.0 \%$ | $6.2 \%$ | $6.2 \%$ | Not defined | $6.2 \%$ |

* Harvest rate (HR).


### 17.7 Management strategies

The EU multiannual plan (MAP) for stocks in the Western Waters and adjacent waters applies to this stock. The plan specifies conditions for setting fishing opportunities depending on stock status and for making use of the FMSY range for the stock. ICES considers the MAP to be precautionary when implemented at the FU level. Full details of the plan are described in EU (2019).

### 17.8 Quality of assessment and forecast

The main quality considerations for this stock are related to mean weight and discarding. The mean weight for this stock has been fluctuating, the most recent estimates maybe overestimate due to the non-inclusion of discards. The mean weight declined from 2014 to 2017 as strong year classes recruit to the fishery. Since 2017 a recent mean weight in the landing was considered the most appropriate basis in the calculation of catch scenarios. In previous years a long-term mean weight was used.

There is good evidence from surveys and length structure of landings that recruitment improved between 2015 and 2017, and this resulted in a reduction in mean weight in the stock in those years. As expected, the mean weight increased in 2018 and 2019 as the stronger cohorts grows;
it decreased in 2020, but increased again in 2021. Currently there is no methodology to take this into account in the calculation of catch options.

Up to 2015 discarding was considered negligible for this functional unit. Since 2016 the amount of discards observed on catch sampling trips have increased. This may be temporary linked to incoming recruitment. Sampling levels are insufficient to estimate total discards accurately, and projections assume no discards. The main concern is that the mean weight derived from the landings grades maybe bias due to unknown discarding levels. Not including discards in the assessment results in an underestimate of the actual fishing pressure. The current estimate is just below FMSY.

The UWTV survey provides abundance since 2012 (except 2015) with high precision, but the time-series is still too short to provide an MSY Btrigger for this FU. The 2022 UWTV survey sampled $88 \%$ of the planned stations; this is considered to have had minimal impact on the abundance estimate and quality of the survey, based on burrow densities in adjoining areas and comparing coefficients of variation from the current and previous survey years.
The landings are considered fairly well estimated up to 2021 (an unallocated component related to area misreporting was included from 2011 to 2017).

### 17.9 Recommendation for next benchmark

This stock was benchmark in 2013 at WKNEPH. WGCSE will keep the stock under close review and recommend future benchmark as required.

### 17.10 Management considerations

There is a separate catch limit for Functional Unit (FU) 16 within the wider TAC for Subarea 7. This has resulted in very restrictive quotas for some vessels which increased area misreporting and the risk of discarding from 2011 to 2017. Area misreporting diminished in 2018 with the introduction of a national legislation restricting Irish vessels' fishing areas, where since March 2018 Irish vessels targeting Nephrops in subareas 6 and 7 may only fish in either of (1) Subarea 6 and Subarea 7, excluding FU16, or (2) FU16 of Sub-area 7 (Fisheries Management Notice No. 20 of 2018). Given the vulnerability of this stock to over exploitation the separate catch limit for Functional Unit (FU) 16 should remain in place.

A seasonal closed area (May 1-July 31) has been in place since 2010. The period of the closure was reduced to one month, May, since 2013. There hasn't been an ICES evaluation of the impact of this closure and whether it provides a conservation benefit over and above catch limits. Some sectors of the fishing industry want to extend the period of closure because they believe that this is a more effective conservation measure than catch limits.

Productivity of deep-water Nephrops stocks is generally lower than that in shelf waters, though individual Nephrops grow to relatively large sizes and attain high market prices. Other deepwater Nephrops stocks off the Spanish and Portuguese coast have collapsed and have been subject to recovery measures for several years e.g. FU25, 26, 27 and 31. Recruitment in Nephrops populations in deep water may be more sporadic than for shelf stocks with strong larval retention mechanisms. This makes these stocks more vulnerable to over exploitation and potential recruitment failure as has been observed on the Porcupine Bank in the early 2000s.
From 2019 vessels using highly selective gears in Subarea 7 can be exempted from the landings obligation on the basis of the high survival exemption (see discard plans). It is unknown if Nephrops discarded on the Porcupine Bank could actually survive the discarding process.

Discarding by the Nephrops trawl fishery is around $50 \%$ of the total catch by weight. The main species that are discarded by weight are blue mouth-red fish, blue whiting and argentines (Marine Institute and Bord Iascaigh Mhara, 2011).

### 17.11 References

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Table 20.1. Nephrops Porcupine Bank (FU 16): Of which catch limit.

| Year | France | Ireland | Spain | UK | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 241 | 454 | 377 | 188 | 1260 |
| 2012 | 238 | 457 | 380 | 185 | 1260 |
| 2013 | 340 | 653 | 543 | 264 | 1800 |
| 2014 | 349 | 671 | 557 | 271 | 1848 |
| 2015 | 349 | 671 | 558 | 272 | 1850 |
| 2016 | 349 | 671 | 558 | 272 | 1850 |
| 2017 | 586 | 1124 | 935 | 455 | 3100 |
| 2018 | 516 | 992 | 825 | 401 | 2734 |
| 2019 | 500 | 959 | 798 | 388 | 2645 |
| 2020 | 498 | 957 | 795 | 387 | 2637 |
| 2021 | 621 | 1194 | 992 | 483 | 3290 |
| 2022 | 530 | 1016 | 846 | 412 | 2804 |

Table 20.2. Nephrops Porcupine Bank (FU 16): Landings (tonnes) by country.

| Year | France | Ireland | Spain | UK (E\& W) | UK (NI) | UK (Scotland) | Unallocated | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 514 |  |  |  |  |  |  | 514 |
| 1966 | 0 |  |  |  |  |  |  | 0 |
| 1967 | 441 |  |  |  |  |  |  | 441 |
| 1968 | 441 |  |  |  |  |  |  | 441 |
| 1969 | 609 |  |  |  |  |  |  | 609 |
| 1970 | 256 |  |  |  |  |  |  | 256 |
| 1971 | 500 |  | 1444 |  |  |  |  | 1944 |
| 1972 | 0 |  | 1738 |  |  |  |  | 1738 |
| 1973 | 811 |  | 2135 |  |  |  |  | 2946 |
| 1974 | 900 |  | 1894 |  |  |  |  | 2794 |
| 1975 | 0 |  | 2150 |  |  |  |  | 2150 |
| 1976 | 6 |  | 1321 |  |  |  |  | 1327 |
| 1977 | 0 |  | 1545 |  |  |  |  | 1545 |
| 1978 | 2 |  | 1742 |  |  |  |  | 1744 |
| 1979 | 14 |  | 2255 |  |  |  |  | 2269 |
| 1980 | 21 |  | 2904 |  |  |  |  | 2925 |
| 1981 | 66 |  | 3315 |  |  |  |  | 3381 |
| 1982 | 358 |  | 3931 |  |  |  |  | 4289 |
| 1983 | 615 |  | 2811 |  |  |  |  | 3426 |
| 1984 | 1067 |  | 2504 |  |  |  |  | 3571 |
| 1985 | 1181 |  | 2738 |  |  |  |  | 3919 |
| 1986 | 1060 |  | 1462 | 69 |  |  |  | 2591 |
| 1987 | 609 |  | 1677 | 213 |  |  |  | 2499 |
| 1988 | 600 |  | 1555 | 220 |  |  |  | 2375 |
| 1989 | 324 | 350 | 1417 | 24 |  |  |  | 2115 |
| 1990 | 336 | 169 | 1349 | 41 |  |  |  | 1895 |
| 1991 | 348 | 170 | 1021 | 101 |  |  |  | 1640 |
| 1992 | 665 | 311 | 822 | 217 |  |  |  | 2015 |
| 1993 | 799 | 206 | 752 | 100 |  |  |  | 1857 |


| Year | France | Ireland | Spain | UK (E\& W) | UK (NI) | UK (Scotland) | Unallocated | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 1088 | 512 | 809 | 103 |  |  |  | 2512 |
| 1995 | 1234 | 971 | 579 | 152 |  |  |  | 2936 |
| 1996 | 1069 | 508 | 471 | 182 |  |  |  | 2230 |
| 1997 | 1028 | 653 | 473 | 255 |  |  |  | 2409 |
| 1998 | 879 | 598 | 405 | 273 |  |  |  | 2155 |
| 1999 | 1047 | 609 | 448 | 185 |  |  |  | 2290 |
| 2000 | 351 | 227 | 213 | 120 |  |  |  | 910 |
| 2001 | 425 | 369 | 270 | 158 |  |  |  | 1222 |
| 2002 | 369 | 543 | 276 | 139 |  |  |  | 1327 |
| 2003 | 131 | 307 | 489 | 108 |  | 29 |  | 1064 |
| 2004 | 289 | 494 | 468 | 126 |  | 28 |  | 1406 |
| 2005 | 397 | 754 | 681 | 208 |  | 156 |  | 2197 |
| 2006 | 462 | 731 | 636 | 201 |  | 155 |  | 2185 |
| 2007 | 302 | 1060 | 384 | 146 |  | 183 |  | 2074 |
| 2008 | 26 | 562 | 234 | 41 |  | 138 |  | 1000 |
| 2009 | 4 | 356 | 348 | 13 |  | 159 |  | 879 |
| 2010 | 4 | 579 | 240 | 10 |  | 90 |  | 922 |
| 2011 | 8 | 643 | 182 | 23 |  | 122 | 301 | 1278 |
| 2012 | 0.46 | 605 | 198 | 0 |  | 134 | 320 | 1258 |
| 2013 | 5.8 | 651 | 132 | 1 |  | 118 | 234 | 1141 |
| 2014 | 3 | 813 | 129 | 0 |  | 96 | 148 | 1189 |
| 2015 | 3 | 744 | 84 | 0 |  | 109 | 454 | 1394 |
| 2016 | 35 | 1052 | 58 | 1 |  | 160 | 849 | 2154 |
| 2017 | 63 | 743 | 73 | 4 | 245 | 131 | 1373 | 2632 |
| 2018 | 81 | 2079 | 158 | 8 | 280 | 144 | 0 | 2751 |
| 2019 | 54 | 1529 | 112 | 7 | 325 | 201 | 0 | 2229 |
| 2020 | 41 | 1516 | 82 | 1 | 259 | <1 | 0 | 1899 |
| 2021 | 49 | 1611 | 318 | 1 | 329 | 169 | 0 | 2476 |

Table 20.3. Nephrops Porcupine Bank (FU 16): Recent sampling used in the assessment.

| Year | Spain | France | Ireland |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
|  | Number of Trips | Type | Number of Trips | Type | Number of Trips | Type | Graded Landings |
| :--- |
| 2010 |
| 2011 |

Table 20.4. Nephrops Porcupine Bank (FU 16): Mean sizes (mm CL) of male and female Nephrops in Spanish, French and Irish landings and the Spanish Porcupine Groundfish survey 1981-2021.

| Year | Spain <br> Landings |  | Ireland <br> Landings |  | France <br> Landings |  | Porcupine Survey <br> Catch |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
|  | Males | Females | Males | Females | Males | Females | Males | Females |
| 1981 | 39.9 | 34.5 | - | - | - | - | - | - |
| 1982 | 40.9 | 34.8 | - | - | - | - | - | - |
| 1983 | 40.8 | 34.0 | - | - | - | - | - | - |
| 1984 | 39.7 | 33.1 | - | - | - | - | - | - |
| 1985 | 38.7 | 33.5 | - | - | - | - | - | - |
| 1986 | 40.7 | 36.4 | - | - | - | - | - | - |
| 1987 | 39.3 | 35.0 | - | - | - | - | - | - |
| 1988 | 40.7 | 38.3 | - | - | - | - | - | - |
| 1989 | 40.5 | 36.8 | - | - | - | - | - | - |
| 1990 | 41.0 | 36.1 | - | - | - | - | - | - |
| 1991 | 39.4 | 34.5 | - | - | - | - | - | - |
| 1992 | 39.2 | 34.1 | - | - | - | - | - | - |
| 1993 | 41.6 | 36.1 | - | - | - | - | - | - |
| 1994 | 40.8 | 36.5 | - | - | - | - | - | - |
| 1995 | 41.3 | 36.6 | 40.7 | 36.5 | 43.2 | 38.3 | - | - |
| 1996 | 41.6 | 35.1 | 34.6 | 35.3 | 41.7 | 38.9 | - | - |
| 1997 | 39.7 | 34.8 | 35.9 | 34.5 | 41.9 | 38.4 | - | - |
| 1998 | 41.1 | 34.6 | 37.2 | 35.6 | 41.9 | 38.4 | - | - |
| 1999 | 41.5 | 35.7 | 36.6 | 33.7 | 43.1 | 39.1 | - | - |
| 2000 | 41.1 | 34.8 | na | na | 45.3 | 40.5 | - | - |
| 2001 | 41.1 | 36.3 | 37.8 | 35.4 | 45.4 | 39.4 | 36.0 | 28.9 |
| 2002 | 39.7 | 35.3 | 36.1 | 38.5 | 45.3 | 40.3 | 37.5 | 31.7 |
| 2003 | 41.4 | 37.8 | 44.5 | 36.2 | 46.2 | 38.9 | 39.7 | 30.9 |
| 2004 | 43.5 | 38.5 | 43.5 | 35.7 | 46.4 | 41.5 | 39.9 | 30.5 |
| 2005 | 43.4 | 38.1 | 46.9 | 40.6 | 45.9 | 41.0 | 45.1 | 33.8 |
| 2006 | 43.9 | 38.0 | na | na | 48.9 | 41.4 | 44.3 | 35.0 |


| Year | Spain <br> Landings |  | Ireland <br> Landings |  | France <br> Landings |  | Porcupine Survey <br> Catch |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
|  | Males | Females | Males | Females | Males | Females | Males | Females |
| 2007 | 43.7 | 41.0 | na | na | 48.3 | 43.8 | 45.9 | 37.8 |
| 2008 | 51.0 | 40.6 | 43.3 | 37.5 | na | na | 48.8 | 38.7 |
| 2009 | 43.0 | 42.7 | 44.1 | 40.1 | na | na | 32.6 | 28.9 |
| 2010 | na | na | 43.2 | 40.4 | na | na | 36.3 | 31.8 |
| 2011 | na | na | 39.5 | 38.4 | na | na | 39.0 | 33.6 |
| 2012 | na | na | 41.1 | 38.1 | na | na | 41.1 | 30.8 |
| 2013 | na | na | 42.9 | 38.9 | na | na | 37.6 | 25.1 |
| 2014 | na | na | 45.1 | 40.9 | na | na | 36.4 | 31.0 |
| 2015 | na | na | 40.3 | 39.7 | na | na | 35.5 | 32.7 |
| 2016 | na | na | 37.8 | 37.3 | na | na | 32.2 | 27.8 |
| 2017 | na | na | 35.7 | 32.9 | na | na | 34.1 | 26.8 |
| 2018 | na | na | 38.8 | 35.3 | na | na | 35.0 | 28.2 |
| 2019 | na | na | 41.3 | 36.2 | na | na | 35.2 | 29.3 |
| 2020 | na | na | na | na | na | na | 37.5 | 29.0 |
| 2021 | na | na | na | na | na | na | 34.6 | 29.8 |

Table 20.5. Nephrops Porcupine Bank (FU16): Time-series of numbers landed and mean weight in the landings.

| Year | Numbers (millions) | Weight Landed (Tonnes) | Mean Weight in landings (gr) |
| :---: | :---: | :---: | :---: |
| 1986 | 55.7 | 2591 | 46.53 |
| 1987 | 60.3 | 2499 | 41.42 |
| 1988 | 48.1 | 2375 | 49.34 |
| 1989 | 45.6 | 2115 | 46.4 |
| 1990 | 38.9 | 1895 | 48.67 |
| 1991 | 37.3 | 1640 | 43.98 |
| 1992 | 47 | 2015 | 42.84 |
| 1993 | 38.5 | 1857 | 48.29 |
| 1994 | 54.4 | 2512 | 46.15 |
| 1995 | 65.5 | 2936 | 44.79 |
| 1996 | 52.9 | 2230 | 42.15 |
| 1997 | 59.1 | 2409 | 40.73 |
| 1998 | 49.9 | 2155 | 43.16 |
| 1999 | 52.3 | 2290 | 43.76 |
| 2000 | 15.1 | 910 | 60.13 |
| 2001 | 24.6 | 1222 | 49.65 |
| 2002 | 32 | 1327 | 41.49 |
| 2003 | 18.4 | 1064 | 57.76 |
| 2004 | 21.5 | 1406 | 65.28 |
| 2005 | 31.5 | 2197 | 69.84 |
| 2006 | 28.7 | 2185 | 76.24 |
| 2007 | 29.2 | 2074 | 71.05 |
| 2008 | 17.9 | 1000 | 55.89 |
| 2009 | 16.5 | 879 | 53.19 |
| 2010 | 14.1 | 922 | 65.32 |
| 2011 | 27.9 | 1278 | 45.81 |
| 2012 | 25.0 | 1258 | 50.36 |
| 2013 | 19.8 | 1141 | 57.54 |
| 2014 | 17.3 | 1189 | 68.54 |


| Year | Numbers (millions) | Weight Landed (Tonnes) | Mean Weight in landings (gr) |
| :--- | :---: | :---: | :---: |
| 2015 | 27.4 | 1394 | 50.86 |
| 2016 | 53.5 | 2154 | 40.29 |
| 2017 | 84.9 | 2632 | 31.01 |
| 2018 | 66.2 | 2751 | 41.55 |
| 2019 | 42.2 | 1899 | 53.38 |
| 2020 | 57.8 | 2476 | 42.26 |
| 2021 | Average 2019-2021 |  | 44.82 |

## Table 20.6. Nephrops Porcupine Bank (FU16): Assessment summary.

|  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\stackrel{\grave{\pi}}{\stackrel{\text { ® }}{\sim}}$ | Millions |  |  |  |  | \% | tonnes |  | \% |  | grammes |  |
| 2012 | 787 | 79 | 25 | 0 | 25 | 3.2 | 1258 | 0 | 0 | 0 | 50.4 | NA |
| 2013 | 768 | 61 | 20 | 0 | 20 | 2.6 | 1141 | 0 | 0 | 0 | 57.5 | NA |
| 2014 | 722 | 35 | 17 | 0 | 17 | 2.4 | 1189 | 0 | 0 | 0 | 68.5 | NA |
| 2015 | NA | NA | 27 | 0 | 27 | 3.3 *** | 1394 | 0 | 0 | 0 | 50.9 | NA |
| 2016 | 958 | 68 | 53 | NA | 53 | 5.6 | 2154 | NA | NA | NA | 40.3 | NA |
| 2017 | 850 | 90 | 85 | NA | 85 | 10.0 | 2632 | NA | NA | NA | 31.0 | NA |
| 2018 | 1117 | 92 | 66 | NA | 66 | 5.9 | 2751 | NA | NA | NA | 41.6 | NA |
| 2019 | 1010 | 101 | 42 | NA | 42 | 4.2 | 2251 | NA | NA | NA | 53.4 | NA |
| 2020 | 1264 | 94 | 50 | NA | 50 | 3.9 | 1899 | NA | NA | NA | 38.3 | NA |
| 2021 | 1018 | 92 | 58 | NA | 58 | 5.7 | 2476 | NA | NA | NA | 42.8 | NA |
| 2022 | 1363 | 91 |  |  |  |  |  |  |  |  |  |  |

*Discarding up to 2015 was considered to be negligible. Discard estimates are not available since 2016 and are therefore not included in the assessment.
** Values since 2016 onwards may be underestimated owing to insufficient discard data.
*** The harvest rate is estimated based on a linear interpolation of abundance, as no survey was carried out in this year. NA = not available.

Table 20.7. Nephrops Porcupine Bank (FU16): Effort and Ipue for the various different fleets exploiting the stock 19712021.

| Year | Spain ${ }^{1}$ |  | France ${ }^{2}$ |  | Ireland ${ }^{3}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort ('000's Hrs) | Lpue (kg/hr) | Effort ${ }^{2}$ ('000's Hrs) | Lpue (>10\%) (kg/hr) | Effort ${ }^{3}$ ('000's KwDays) | Lpue (t/KWdays) |
| 1980 | 318 | 9 |  |  |  |  |
| 1981 | 272 | 12 |  |  |  |  |
| 1982 | 237 | 17 |  |  |  |  |
| 1983 | 196 | 14 | 18 | 35 |  |  |
| 1984 | 194 | 13 | 30 | 35 |  |  |
| 1985 | 200 | 14 | 33 | 36 |  |  |
| 1986 | 162 | 9 | 28 | 38 |  |  |
| 1987 | 174 | 10 | 24 | 26 |  |  |
| 1988 | 180 | 9 | 22 | 27 |  |  |
| 1989 | 173 | 8 | 14 | 23 |  |  |
| 1990 | 159 | 9 | 15 | 23 |  |  |
| 1991 | 138 | 7 | 19 | 18 |  |  |
| 1992 | 96 | 9 | 32 | 21 |  |  |
| 1993 | 80 | 9 | 36 | 22 |  |  |
| 1994 | 80 | 10 | 38 | 28 |  |  |
| 1995 | 67 | 9 | 42 | 30 | 584.9 | 1.4 |
| 1996 | 58 | 8 | 41 | 26 | 192.5 | 1.59 |
| 1997 | 57 | 8 | 41 | 25 | 327.3 | 1.26 |
| 1998 | 56 | 7 | 40 | 22 | 284.6 | 1.59 |
| 1999 | 53 | 8 | 43 | 21 | 278 | 1.29 |
| 2000 | 47 | 5 | 23 | 14 | 92.8 | 1.25 |


| Year | Spain ${ }^{1}$ |  | France ${ }^{2}$ |  | Ireland ${ }^{3}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort ('000's Hrs) | Lpue (kg/hr) | Effort ${ }^{2}$ ('000's Hrs) | Lpue (>10\%) (kg/hr) | Effort ${ }^{3}$ ('000's KwDays) | Lpue (t/KWdays) |
| 2001 | 44 | 6 | 24 | 15 | 230.2 | 1.12 |
| 2002 | 54 | 5 | 18 | 18 | 339.8 | 1.3 |
| 2003 | 66 | 5 | 7 | 19 | 294.7 | 0.8 |
| 2004 | 59 | 10 | 9 | 25 | 569.2 | 0.68 |
| 2005 | 60 | 13 | 15 | 26 | 756.2 | 0.83 |
| 2006 | 65 | 9 | 22 | 21 | 952.8 | 0.72 |
| 2007 | 58 | 8 | 17 | 18 | 1199.4 | 0.81 |
| 2008 | 42 | 6 | 4 | 7 | 830.7 | 0.67 |
| 2009 | 44 | 7 | na | na | 411.3 | 0.83 |
| 2010 | 42 | 6 | na | na | 704.1 | 0.81 |
| 2011 | na | na | na | na | 986.9 | 0.63 |
| 2012 | 15 | na | na | na | 817.1 | 0.63 |
| 2013 | na | na | na | na | 885.7 | 0.92 |
| 2014 | na | na | na | na | 1019.8 | 0.92 |
| 2015 | na | na | na | na | 1219.2 | 0.99 |
| 2016 | na | na | na | na | 1359.3 | 1.43 |
| 2017 | na | na | na | na | 1328.9 | 1.59 |
| 2018 | na | na | na | na | 1721.2 | 1.21 |
| 2019 | na | na | na | na | 1463.3 | 1.01 |
| 2020 | na | na | na | na | 1468.3 | 0.97 |
| 2021 | na | na | na | na | 1647.1 | 0.94 |

${ }^{1}=$ Effort and lpue between 1980 and 2010 was estimated based on fishing days in 7 . Effort in 2012 was based on logbooks for FU16.
$2=$ Effort and lpue for vessels where $<\mathbf{1 0 \%}$ of landed value was Nephrops.
3 = Effort and lpue for vessels where $30 \%$ of the landed weight was Nephrops.


Figure 20.1. Nephrops in FU16 (Porcupine Bank). WG's best estimates of landings in tonnes by country.


Figure 20.2. Nephrops in FU16 (Porcupine Bank). Female and male length distributions of raised international landings. Vertical dashed lines refer to Minimum Landing Size ( 25 mm ).


Figure 20.3. Nephrops in FU16 (Porcupine Bank). The percentage males in the landings and survey over time.


Figure 20.4. Nephrops in FU16 (Porcupine Bank). Mean weight in the commercial landings.

Catches ( $\mathbf{k g} \times 30 \mathrm{~min}$ haul-1)


Number of juveniles ( $\mathbf{\leq} \mathbf{2 0} \mathbf{~ m m}$ )


Figure 20.5. Nephrops in FU16 (Porcupine Bank). Geographic distribution of Nephrops norvegicus in Porcupine surveys between 2010 and 2021. Top panel: catches ( $\mathbf{k g \times 3 0} \mathbf{~ m i n}$ haul-1). Bottom panel: Number of juveniles $(\leq 20 \mathrm{~mm}$ carapace length). (Velasco et al. in draft)



Year

Figure 20.6. Nephrops in FU16 (Porcupine Bank). Evolution of Nephrops norvegicus biomass and abundance indices in Porcupine surveys (2001-2021). Boxes mark parametric standard error of the stratified abundance index. Lines mark boot-strap confidence intervals ( $\alpha=0.80$, bootstrap iterations = 1000). (Velasco et al. in draft)

 and 3+ Nephrops (>33 mm) in Porcupine survey 2001-2021. (Velasco et al. in draft)

## Landings



Fishing pressure


Figure 20.8. Nephrops in FU16 (Porcupine Bank). Summary of stock status for Porcupine Nephrops.

## 18 Norway lobster (Nephrops norvegicus) in Division 7.b, Functional Unit 17 (west of Ireland, Aran grounds)

## Type of assessment in 2022

This stock was inter-benchmarked in September 2015 by correspondence (ICES, 2016a). The assessment and catch options follow the agreed procedures set out in the stock annex.

## ICES advice applicable to 2021

"ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, catches in 2021 that correspond to the F ranges in the MAP are between 443 tonnes and 508 tonnes, assuming recent discard rates. The entire range is considered precautionary when applying the ICES advice rule.

To ensure that the stock in Functional Unit (FU) 17 is exploited sustainably, management should be implemented at the functional unit level."

## ICES advice applicable to 2022

"ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2018-2020, catches in 2022 that correspond to the F ranges in the MAP are between 313 and 360 tonnes.

To ensure that the stock in Functional Unit (FU) 17 is exploited sustainably, management should be implemented at the FU level."

### 18.1 General

## Stock description and management units

The Aran Grounds Nephrops stock (FU17) covers ICES rectangles 34-35 D9-E0 within 7.b. This stock is included as part of the TAC Area 7 Nephrops which includes the following stocks: Irish Sea East and West (FU14, FU15), Porcupine Bank (FU16), northwestern Irish Coast (FU18), southeastern and south-western Irish Coast (FU19) and the Celtic Sea (FU20-22).

Map below shows FU17 assessment area (blue) and TAC area (red). See Section 18 for details on Nephrops Subarea 7 general section.


## Ecosystem aspects

Details of the ecosystem on the Aran grounds are provided in the stock annex updated by IBPNeph (ICES, 2016a).

## Fishery description

A description of the fleet is given in the stock annex. The time-series of numbers of vessels is updated in Figure 21.1.1. The numbers of vessels had been relatively stable from 1995 to 2018, but it decreased in 2019 and 2020 to half, and increased again to previous numbers in 2021. The time-series of vessel power is shown as a box and kite plot in Figure 21.1.2.

The majority of the landings are made with 80 mm mesh.
The majority of the landings come from the grounds to the west and southwest of the Aran Islands known as the 'back of the Aran ground' (See stock annex). The fishery on the Aran Grounds operates throughout the year, weather permitting with a seasonal trend (See stock annex).

## Fishery in 2021

In recent years several newer vessels specializing in Nephrops fishing have participated periodically in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates. There has been a trend for Irish vessels to switch to multi (quad) rig trawls since 2012. These vessels are more efficient at catching Nephrops (BIM, 2014).

## Information from stakeholders

Voluntary effort restriction were put in place by the Irish fishing industry in April and May 2015. These measures reduced catches and effort significantly on the stock in advance of the 2015 UWTV survey.

### 18.2 Data

## InterCatch

Data were available in InterCatch and used for catch data only.

## Landings

The reported landings time-series is shown in Figure 21.2.1 and Table 21.2.1. The 2021 landings increased by about $124 \%$ from those made in 2020 and amounted to 498 t .

## Effort

The IBPNeph 2015 reviewed Irish commercial landings and effort data in detail. They concluded that effort should be reported in the WGCSE report in KWdays and lpue should be reported in $\mathrm{KG} / \mathrm{kwdays}$ in the knowledge that the trend is likely to be a biased underestimate because it is not adjusted for efficiency or behavioural changes. The time-series of effort and lpue is updated in Figure 21.2.2 and Table 21.2.2. There was a significant decline in lpue and effort in 2015 due to the local management efforts put in place in April and May. In 2016 effort level increased to values similar to those observed previously prior to 2011. However, since 2017 effort levels have declined, reaching in 2020 the lowest values in the data series, and increasing slightly in 2021.

## Sampling levels

Sampling levels, data aggregating and raising procedures were reviewed by IBPNeph 2015 and are documented in the stock annex. The time-series of samples is shown in Figure 21.2.3 and Table 21.2.3. Given the low level of landings in recent years, it has been challenging to obtain sufficient samples to provide robust estimates of mean weights.

## Commercial length-frequency distributions

The raised catch length distributions are shown in Figure 21.2.4. The mean length of females decreased in 2018, increasing the discard rate for females. Female lengths increased in 2019 and 2020 to similar values of 2017, and they decreased again in 2021, but not to the levels of 2018.

## Sex ratio

In 2021 the difference on the proportion of males between the catches and the landings is similar to values from 2017 to 2019 (Figure 21.2.5). Sex ratio has a distinct seasonal pattern with lowest male proportions in the samples in May and June. Males dominate the catches in the autumn and winter.

## Mean weight explorations

Explorations of the mean weight in the catch samples by sex shows a strong cyclical pattern in females, which corresponds with the emergence of mature females from the burrows to mate in summer (Figure 21.2.6). The annual mean weight estimate for landings and discards is shown in Figure 21.2.7. The mean weight estimates have been relatively stable from 2011, where main
change occurred in 2008-2011. In 2020 the discards mean weight increased significantly, but in 2021 it decreased to similar values of previous years.

## Discarding

Table 21.2.4 gives weights, numbers and proportions of the landings and discard raised internationally according to the stock annex. A $25 \%$ discard survival rate is assumed in line with other Nephrops stocks in the Celtic sea (see stock annex) as the basis for the catch scenarios. Gear selectivity trials by Bord Iascaigh Mhara (BIM, 2017) reported a $64 \%$ survivor rate for Nephrops caught in a trawl with a SELTRA selectivity device in the outer Galway Bay area.

## Abundance indices from UWTV surveys

The spatial extent of the Nephrops grounds in FU17 was re-defined by IBPNeph 2015 and the total abundance estimates were revised using a new procedure (ICES, 2016a). The redefinition of the polygons in FU17 resulted in $\sim 30 \%$ increase in overall area from $1007 \mathrm{~km}^{2}$ to $1320 \mathrm{~km}^{2}$ (stock annex). The 2022 UWTV survey was not deemed robust enough for the assessment because of the reduced number of stations completed due to disruption to the survey schedule (Aristegui et al., 2022). As such, the stock size is unknown for 2022. The assessment and advice is therefore based on the 2021 UWTV survey.

The spatial distributions of burrow densities are shown in Figure 21.2.8. The densities have fluctuated considerably over the time-series and throughout the Aran grounds. In general, the densities are higher towards the middle-western side of the ground and there is a notable trend towards lower densities towards the east. On the south-western boundary, there are often high densities close to the boundary. In this area, there is a sharp transition from mud to rocky substrate.

The summary statistics from this geostatistical analysis are given in Table 21.2.5 and plotted in Figure 21.2.10. The geostatistical abundance estimate adjusted is derived using the mean of the krigged grid where the mean of the observations is reported in Table 21.2.5. In recent years the Aran Grounds accounted for $\sim 92 \%$ of the total estimated burrow abundance from FU17 (Table 21.2.5). Galway Bay accounted for $\sim 5 \%$ and Slyne Head for $\sim 3 \%$ (Table 21.2.6). The Galway Bay estimates fluctuate widely but are highly correlated with the Aran ground (Figure 21.2.9). Estimates for the Slyne Head ground also fluctuate considerably but show no significant correlation with the other areas except for the peaks of 2010, 2015 and 2018 (Figure 21.2.9).

Aran ground abundance estimate's CV (Table 21.2.5) has been always well below the recommendation of $20 \%$ by SGNEPS (ICES, 2012). The CV on the abundance estimates for Galway Bay and Slyne Head have also stayed low (Table 21.2.6) and within the recommendation, showing the surveys are precise. Figure 21.2.10 and Table 21.2.7 show the total abundance estimate for FU17 with the IBPNeph proposed MSY B ${ }_{\text {trigger. }}$. The 2022 abundance estimate is based on the 2021 UWTV survey, and thus has not changed from last year's 331 million, which is below the MSY $B_{\text {trigger }}$ ( 540 million).

### 18.3 Assessment

## Comparison with previous assessments

The WGCSE 2022 carried out an UWTV based assessment for this stock. The methods used were very much in line with WKNEPH (ICES, 2009a) and the approach taken for other Nephrops stocks in 6 and 7 by WGCSE. This approach was inter-benchmarked at IBPNeph (ICES, 2016a).

## State of the stock

The stock size is unknown for 2022. The assessment and advice is therefore based on the 2021 UWTV survey. UWTV abundance estimates suggest that the stock size has fluctuated widely with an overall declining trend and is below MSY Btrigger since 2012 (except 2015 and 2018). The 2021 estimate was lowest observed in the time-series and was below the MSY $B_{\text {trigger. }}$. The 2021 abundance remained below the average of the series (geomean [2002-2019]: 617 million). Harvest rate [calculated as (landings + dead discards)/abundance estimate] was below the FmsYproxy from 2017 to 2020, and increased up to Fmšproxy in 2021 (Table 21.3.1 and Figure 21.3.1).

### 18.4 Catch scenario table

Catch scenario table inputs and historical estimates of mean weight in landings and harvest ratios are presented in Table 21.3.1 and summarised below. The calculation of catch options for the Aran Grounds follows the procedure outlined in the stock annex.

The basis for the catch scenarios.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Stock abundance (2023) | 331 | UWTV survey 2021; numbers of individuals in millions |
| Mean weight in projected landings | 22.4 | Average 2008-2021; in grammes |
| Mean weight in projected discards | 11.4 | Average 2008-2021; in grammes |
| Projected discards | 24.1 | Average 2019-2021; percentage by number |
| Discards survival | 25 | Percentage by number |

Given the fluctuations observed in mean weights for landings and discards (Figure 21.2.7) an average from 2008 to the most recent year is used in the calculation of catch options as set out in the stock annex. The discard rates and proportions for the last three years are used to account for recent on-board retention practices (this is also according to the stock annex).

### 18.5 Reference points

New reference points were defined for this stock at the IBPNeph (ICES, 2016a) and no new proposals were made by WKMSYRef4 (ICES, 2016b). For Nephrops stocks MSY Btrigger has been defined as the lowest stock size from which the abundance has increased. This corresponds to the abundance observed in 2008 rounded to the nearest $10=540$ million individuals (Figure 21.2.10 and Table 21.2.7).

The Fmsy proxy was revised during the benchmark in 2015. The observed burrow density has declined, from high ( $>0.8$ individuals $\mathrm{m}^{-2}$ ) at the start of the series to medium density ( $\sim 0.3$ individuals $\mathrm{m}^{-2}$ ) towards the end of the time-series. The nature of the fishery has also changed, from a continuous fishery throughout the year to a fishery which is more concentrated on sporadic periods of high catch rates. For these reasons a harvest rate consistent with a combined sex $\mathrm{F}_{0.1}=$ $8.5 \%$ is considered an appropriate proxy for $\mathrm{F}_{\text {MSY. }}$

These should remain under review by WGCSE and may be revised while data become available.

### 18.6 Management strategies

As yet there are no explicit management strategies for this stock but there have been some discussions among the fishing industry and scientists about developing a long-term plan for the management of the Aran fishery. Sustainable utilization of the Nephrops stock will form the cornerstone of any management strategy for this fishery.

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to Norway lobster (Nephrops norvegicus) by functional unit in ICES Subarea 7 and also demersal stocks.

### 18.7 Quality of assessment and forecast

Biological sampling for this stock is adequate. From 2002 to 2021 a dedicated annual UWTV survey has provided abundance estimates for the Aran Grounds with high precision. However, in 2022 the UWTV survey was not deemed robust enough for the assessment because of the reduced number of stations completed due to disruption to the survey schedule. The area of the Aran Grounds was revised in 2015, resulting in a recalculation of the abundance time-series which now also includes Galway Bay and Slyne Head. A number of other biological parameters such as mean weights and length distributions have also been revised. The revisions were made as part of an inter-benchmark process and have improved the quality of the assessment.

In the provision of catch options based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. From 2016, fisheries catching Nephrops in Subarea 7 are covered by the EU landings obligation (EU, 2015). Creel fisheries are exempted from the landings obligation, with a de minimis exemption consisting of a $5 \%$ discard rate by weight for the trawl fishery in 2019 (reduced from 6\% in 2018 and 7\% in both 2016 and 2017). The average discard rate by weight for FU17 over the last three years is $12.4 \%$. Catch advice and scenarios are provided this year on the assumption that discarding is assumed to continue at recent average.

Irish discard survival experiments indicate that the trawl discard survival may be around $64 \%$ (BIM, 2017). As a result, an exemption from the landings obligation based on high survivability has been granted by the European Commission. ICES continues to use the survival rate of $25 \%$ (ICES, 2016c) as the survival rates estimated by BIM (2017) have not been evaluated by ICES.

There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNeph 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008; SGNEPS 2009b; WGNEPS 2014; WKNEPS 2016d; Dobby et al., 2021). Ultimately there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate, although no more precise (ICES, 2009a).

Landings data were adjusted to take into account landings that had been misreported from FU16 from 2011 to 2017. This adjustment is thought to be reasonably accurate (See Section 18).

### 18.8 Recommendation for next benchmark

This stock was last benchmarked by IBPNeph (ICES, 2016a). WGCSE will keep the stock under close review and recommend future benchmark as required.

### 18.9 Management considerations

A meeting was held with stakeholders in March 2015 to discuss the state of the Aran Nephrops stock. In response to this meeting voluntary effort limits were put in place for April, May and June 2015. These voluntary measures have significantly reduced effort and catches on the Aran grounds in 2015 before the UWTV survey.

Small whole Nephrops are the main species comprising the discards. The main fish species discarded are haddock, hake, whiting, megrim and dogfish (Anon, 2011).

The ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management at the Functional Unit level could provide controls to ensure effort and catch were in line with resources available.

### 18.10 References

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Table 21.2.1. Nephrops in FU17 (Aran Grounds). Landings in tonnes by country.

| Year | France | Rep. of Ireland | UK | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1974 | 477 |  |  | 477 |
| 1975 | 822 |  |  | 822 |
| 1976 | 131 |  |  | 131 |
| 1977 | 272 |  |  | 272 |
| 1978 | 481 |  |  | 481 |
| 1979 | 452 |  |  | 452 |
| 1980 | 442 |  |  | 442 |
| 1981 | 414 |  |  | 414 |
| 1982 | 210 |  |  | 210 |
| 1983 | 131 |  |  | 131 |
| 1984 | 324 |  |  | 324 |
| 1985 | 207 |  |  | 207 |
| 1986 | 147 |  | 1 | 148 |
| 1987 | 62 |  | 0 | 62 |
| 1988 | 14 | 814 |  | 828 |
| 1989 | 27 | 317 | 3 | 347 |
| 1990 | 30 | 489 |  | 519 |
| 1991 | 11 | 399 |  | 410 |
| 1992 | 11 | 361 | 2 | 374 |
| 1993 | 11 | 361 | 0 | 372 |
| 1994 | 18 | 707 | 4 | 729 |
| 1995 | 91 | 774 | 2 | 867 |
| 1996 | 2 | 519 | 7 | 528 |
| 1997 | 2 | 839 | 0 | 841 |
| 1998 | 9 | 1401 | 0 | 1410 |
| 1999 | 0 | 1140 | 0 | 1140 |
| 2000 | 1 | 879 | 0 | 880 |
| 2001 | 1 | 912 | 0 | 913 |
| 2002 | 2 | 1152 | 0 | 1154 |


| Year | France | Rep. of Ireland | UK | Total |
| :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 933 | 0 | 933 |
| 2004 | 0 | 525 | 0 | 525 |
| 2005 | 0 | 778 | 0 | 778 |
| 2006 | 0 | 637 | 0 | 637 |
| 2007 | 0 | 913 | 0 | 913 |
| 2008 | 0 | 1050 | 7 | 1057 |
| 2009 | 0 | 625 | 0 | 625 |
| 2010 | 0 | 930 | 9 | 939 |
| 2011 | 0 | 659 | 0 | 659 |
| 2012 | 0 | 1246 | 0 | 1246 |
| 2013 | 0 | 1295 | 0 | 1295 |
| 2014 | 0 | 766 | 0 | 766 |
| 2015 | 0 | 370 | 0 | 370 |
| 2016 | 0 | 641 | 0 | 641 |
| 2017 | 0 | 295 | 0.4 | 295 |
| 2018 | 0 | 494 | 42 | 536 |
| 2019 | 0 | 162 | 4 | 167 |
| 2020 | 0 | 188 | 34 | 222 |
| 2021 | 0 | 490 | 8 | 498 |

Table 21.2.2. Nephrops in FU17 (Aran Grounds). Effort data for the Irish otter trawl Nephrops directed fleet.

| Year | Effort (Kw Days) | Landings (Kgs) |
| :---: | :---: | :---: |
| 1995 | 286,939 | 522,007 |
| 1996 | 174,030 | 312,421 |
| 1997 | 260,676 | 442,218 |
| 1998 | 445,308 | 940,902 |
| 1999 | 366,839 | 782,407 |
| 2000 | 293,684 | 561,244 |
| 2001 | 362,754 | 586,462 |
| 2002 | 350,346 | 798,744 |
| 2003 | 492,284 | 801,813 |
| 2004 | 355,673 | 420,652 |
| 2005 | 396,202 | 708,540 |
| 2006 | 337,503 | 618,515 |
| 2007 | 460,396 | 905,282 |
| 2008 | 512,245 | 1,052,077 |
| 2009 | 319,873 | 613,220 |
| 2010 | 441,080 | 910,346 |
| 2011 | 332,300 | 667,564 |
| 2012 | 488,721 | 1,139,413 |
| 2013 | 571,916 | 1,239,469 |
| 2014 | 460,818 | 774,097 |
| 2015 | 232,190 | 461,409 |
| 2016 | 396,502 | 578,420 |
| 2017 | 277,117 | 258,052 |
| 2018 | 233,793 | 483,723 |
| 2019 | 136,278 | 148,795 |
| 2020 | 91,263 | 177,895 |
| 2021 | 184,881 | 444,551 |

Table 21.2.3. Nephrops in FU17 (Aran Grounds). Sampling levels.

| Year | Quarter | Number of samples |  | Numbers Measured |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Catch | Discards |
| 2008 | 1 | 2 | 3 | 565 | 1376 |
| 2008 | 2 | 9 | 8 | 2224 | 3758 |
| 2008 | 3 | 5 | 4 | 1266 | 1834 |
| 2008 | 4 | 3 | 3 | 889 | 1733 |
| 2009 | 1 | 3 | 3 | 800 | 1184 |
| 2009 | 2 | 6 | 6 | 1685 | 1978 |
| 2009 | 3 | 6 | 6 | 2260 | 2726 |
| 2009 | 4 | 2 | 2 | 1491 | 1149 |
| 2010 | 1 | 4 | 4 | 3322 | 2322 |
| 2010 | 2 | 8 | 7 | 3577 | 2957 |
| 2010 | 3 | 2 | 2 | 951 | 742 |
| 2010 | 4 | 6 | 4 | 3209 | 1802 |
| 2011 | 1 | 7 | 7 | 3755 | 3537 |
| 2011 | 2 | 7 | 7 | 7399 | 6617 |
| 2011 | 3 | 4 | 2 | 3531 | 2386 |
| 2011 | 4 | 5 | 5 | 2440 | 2271 |
| 2012 | 1 | 3 | 3 | 1538 | 1250 |
| 2012 | 2 | 17 | 15 | 6481 | 5113 |
| 2012 | 3 | 0 | 0 | - | - |
| 2012 | 4 | 5 | 5 | 2333 | 1945 |
| 2013 | 1 | 10 | 9 | 3108 | 2983 |
| 2013 | 2 | 11 | 11 | 3733 | 3733 |
| 2013 | 2 | 3 | 3 | 1163 | 1263 |
| 2013 | 4 | 7 | 7 | 2956 | 1779 |
| 2014 | 1 | 3 | 3 | 1208 | 1223 |
| 2014 | 2 | 12 | 12 | 5365 | 3563 |
| 2014 | 3 | 2 | 2 | 786 | 499 |
| 2014 | 4 | 8 | 8 | 3542 | 2760 |


| Year | Quarter | Number of samples |  | Numbers Measured |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Catch | Discards |
| 2015 | 1 | 2 | 2 | 827 | 611 |
| 2015 | 2 | 2 | 2 | 961 | 664 |
| 2015 | 3 | 0 | 0 | - | - |
| 2015 | 4 | 2 | 2 | 1047 | 1388 |
| 2016 | 1 | 5 | 4 | 2292 | 876 |
| 2016 | 2 | 11 | 11 | 4756 | 3383 |
| 2016 | 3 | 6 | 5 | 3020 | 2048 |
| 2016 | 4 | 6 | 6 | 1389 | 1311 |
| 2017 | 1 | 3 | 3 | 1214 | 845 |
| 2017 | 2 | 6 | 4 | 2911 | 1569 |
| 2017 | 3 | 2 | 1 | 1018 | 223 |
| 2017 | 4 | 3 | 3 | 1176 | 839 |
| 2018 | 1 | 3 | 3 | 1224 | 1241 |
| 2018 | 2 | 8 | 8 | 3179 | 2971 |
| 2018 | 3 | 1 | 1 | 467 | 388 |
| 2018 | 4 | 6 | 6 | 1894 | 2487 |
| 2019 | 1 | 3 | 3 | 1151 | 1368 |
| 2019 | 2 | 5 | 5 | 1552 | 1441 |
| 2019 | 3 | 2 | 2 | 628 | 480 |
| 2019 | 4 | 2 | 2 | 519 | 558 |
| 2020 | 1 | 4 | 4 | 1037 | 984 |
| 2020 | 2 | 5 | 5 | 1706 | 1666 |
| 2020 | 4 | 1 | 0 | 302 | 0 |
| 2021 | 1 | 3 | 3 | 740 | 778 |
| 2021 | 2 | 5 | 4 | 1558 | 1059 |

Table 21.2.4. Nephrops in FU17 (Aran Grounds). Raised landings and discard weight and numbers by year.

| Year | Landings (t) | Discards (t) | Landings in number ('000s) | Discards in number ('000s) | Discards by weight (\%) | Discards by number (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 1057 | 248 | 48,162 | 22,074 | 19.0 | 31.4 |
| 2009 | 626 | 129 | 24,935 | 9,487 | 17.1 | 27.6 |
| 2010 | 939 | 224 | 37,341 | 15,246 | 19.3 | 29.0 |
| 2011 | 659 | 92 | 31,950 | 8,542 | 12.2 | 21.1 |
| 2012 | 1246 | 86 | 61,076 | 8,292 | 6.5 | 12.0 |
| 2013 | 1295 | 129 | 60,016 | 12,034 | 9.1 | 16.7 |
| 2014 | 766 | 48 | 33,882 | 5,038 | 5.9 | 12.9 |
| 2015 | 370 | 15 | 17,693 | 1,622 | 3.8 | 8.4 |
| 2016 | 641 | 69 | 30,231 | 6,375 | 9.7 | 17.4 |
| 2017 | 295 | 38 | 13,269 | 3,605 | 11.3 | 21.4 |
| 2018 | 536 | 106 | 22,049 | 10,490 | 16.5 | 32.2 |
| 2019 | 167 | 21 | 7,568 | 2,098 | 11.1 | 21.7 |
| 2020 | 222 | 54 | 9,516 | 3,525 | 19.5 | 27.0 |
| 2021 | 498 | 88 | 22,750 | 7,049 | 15.0 | 23.6 |

Table 21.2.5. Nephrops in FU17 (Aran Grounds). Results summary table for geostatistical analysis of UWTV survey.

| Ground | Year | Number of stations | Mean Density adjusted** (burrow/m ${ }^{2}$ ) | Domain <br> Area ( $\mathrm{km}^{2}$ ) | Geostatistical Abundance Estimate adjusted (millions burrows) | CV on Burrow estimate \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Aran Grounds | 2002 | 49 | 0.79 | 1196 | 947 | 3 |
|  | 2003 | 41 | 0.94 | 1196 | 1118 | 6 |
|  | 2004 | 64 | 1.08 | 1196 | 1297 | 3 |
|  | 2005 | 70 | 0.81 | 1196 | 972 | 2 |
|  | 2006 | 67 | 0.46 | 1196 | 556 | 3 |
|  | 2007 | 71 | 0.69 | 1196 | 828 | 2 |
|  | 2008 | 63 | 0.41 | 1196 | 494 | 3 |
|  | 2009 | 82 | 0.52 | 1196 | 627 | 2 |
|  | 2010 | 87 | 0.63 | 1196 | 752 | 2 |
|  | 2011 | 76 | 0.51 | 1196 | 609 | 2 |
|  | 2012 | 31* | 0.33 | 1196 | 397 | 3 |
|  | 2013 | 31* | 0.33 | 1196 | 390 | 4 |
|  | 2014 | 33* | 0.28 | 1196 | 332 | 4 |
|  | 2015 | 34* | 0.40 | 1197 | 480 | 4 |
|  | 2016 | 34* | 0.29 | 1197 | 343 | 3 |
|  | 2017 | 31* | 0.31 | 1196 | 377 | 3 |
|  | 2018 | 33* | 0.40 | 1196 | 488 | 3 |
|  | 2019 | 31* | 0.39 | 1196 | 458 | 4 |
|  | 2020 | 34 | 0.29 | 1196 | 359 | 4 |
|  | 2021 | 34* | 0.26 | 1196 | 311 | 4 |

[^11]Table 21.2.6. Nephrops in FU17 (Galway Bay and Slyne Head). Results summary table for analysis of UWTV survey. Random stratified estimates given for these grounds only.

| Ground | Year | Number of stations | Mean Density adjusted (burrow/m²) | Domain Area (km²) | Raised Abundance Estimate adjusted (millions burrows)* | CV on Burrow estimate \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GalwayBay | 2002 | 7 | 1.18 | 79.0 | 93.1 | 7 |
|  | 2003 | 3 | 1.30 | 79.0 | 102.6 | 16 |
|  | 2004 | 8 | 1.17 | 79.0 | 92.2 | 14 |
|  | 2005 | 4 | 1.30 | 79.0 | 103.0 | 11 |
|  | 2006 | 3 | 0.74 | 79.0 | 58.8 | 9 |
|  | 2007 | 5 | 0.91 | 79.0 | 71.8 | 8 |
|  | 2008 | 5 | 0.40 | 79.0 | 31.6 | 4 |
|  | 2009 | 8 | 0.71 | 79.0 | 56.3 | 4 |
|  | 2010 | 10 | 1.24 | 79.0 | 97.6 | 11 |
|  | 2011 | 6 | 0.55 | 79.0 | 43.2 | 12 |
|  | 2012 | 4 | 0.64 | 79.0 | 50.9 | 10 |
|  | 2013 | 5 | 0.37 | 79.0 | 29.6 | 10 |
|  | 2014 | 3 | 0.50 | 79.0 | 39.8 | 6 |
|  | 2015 | 5 | 0.71 | 79.0 | 55.8 | 15 |
|  | 2016 | 7 | 0.32 | 79.0 | 25.1 | 7 |
|  | 2017 | 5 | 0.20 | 79.0 | 15.8 | 4 |
|  | 2018 | 5 | 0.41 | 79.0 | 32.5 | 17 |
|  | 2019 | 5 | 0.29 | 79.0 | 22.8 | 11 |
|  | 2020 | 5 | 0.34 | 79.0 | 27.2 | 13 |
|  | 2021 | 5 | 0.15 | 79.0 | 11.5 | 2 |
|  | 2022* | 5 | 0.19 | 79.0 | 14.8 | 3 |

* 2022 abundance estimate for Galway Bay was not used in the assessment, as the rest of the stations in the Aran Grounds and Slyne Head were not completed due to disruption to the survey schedule.

| Ground | Year | Number of stations | Mean Density adjusted (burrow/m²) | Domain Area (km²) | Raised Abundance Estimate adjusted (millions burrows)* | CV on Burrow estimate \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Slyne <br> Head | 2002 | 5 | 0.76 | 39.1 | 29.8 | 8 |
|  | 2003* | 0 | 0.65 | 39.1 | 25.3 | 0 |
|  | 2004 | 3 | 0.53 | 39.1 | 20.8 | 10 |
|  | 2005 | 3 | 0.44 | 39.1 | 17.4 | 1 |
|  | 2006 | 3 | 0.30 | 39.1 | 11.8 | 9 |
|  | 2007 | 4 | 0.51 | 39.1 | 19.8 | 12 |
|  | 2008* | 0 | 0.41 | 39.1 | 16.0 | 0 |
|  | 2009 | 6 | 0.31 | 39.1 | 12.2 | 7 |
|  | 2010 | 7 | 0.73 | 39.1 | 28.7 | 4 |
|  | 2011 | 7 | 0.51 | 39.1 | 20.0 | 5 |
|  | 2012 | 3 | 0.52 | 39.1 | 20.5 | 2 |
|  | 2013 | 4 | 0.54 | 39.1 | 21.1 | 10 |
|  | 2014 | 4 | 0.28 | 39.1 | 11.0 | 6 |
|  | 2015 | 5 | 0.50 | 39.1 | 19.6 | 4 |
|  | 2016 | 4 | 0.27 | 39.1 | 10.8 | 3 |
|  | 2017 | 4 | 0.27 | 39.1 | 10.7 | 4 |
|  | 2018 | 5 | 0.84 | 39.1 | 33.0 | 12 |
|  | 2019 | 5 | 0.29 | 39.1 | 11.5 | 8 |
|  | 2020 | 5 | 0.19 | 39.1 | 7.4 | 4 |
|  | 2021 | 5 | 0.23 | 39.1 | 9.1 | 2 |

*estimated as no survey data available for these years.

Table 21.2.7. Nephrops in FU17. Results summary table for analysis of UWTV survey for the combined grounds.

| Year | Abundance (Millions) | Upper bound | Lower bound |
| :---: | :---: | :---: | :---: |
| 2002 | 1070 | 1154 | 985 |
| 2003 | 1246 | 1434 | 1059 |
| 2004 | 1410 | 1517 | 1302 |
| 2005 | 1092 | 1154 | 1030 |
| 2006 | 627 | 703 | 551 |
| 2007 | 920 | 982 | 858 |
| 2008 | 541 | 588 | 494 |
| 2009 | 696 | 739 | 653 |
| 2010 | 879 | 926 | 831 |
| 2011 | 672 | 720 | 624 |
| 2012 | 468 | 520 | 417 |
| 2013 | 441 | 506 | 376 |
| 2014 | 383 | 440 | 327 |
| 2015 | 556 | 627 | 484 |
| 2016 | 379 | 420 | 339 |
| 2017 | 404 | 445 | 362 |
| 2018 | 554 | 637 | 471 |
| 2019 | 493 | 558 | 427 |
| 2020 | 394 | 453 | 335 |
| 2021 | 331 | 362 | 301 |
| 2022 | 331* |  |  |

[^12]Table 21.3.1. Nephrops in FU17 (Aran Grounds). Forecast inputs (bold) and historical estimates of mean weight in landings and harvest rate. Removals estimated in years with no sampling (*) using ratio of removals to landings in adjacent years. $\mathrm{n} / \mathrm{a}=$ not available due to non-cooperation with sampling programmes.

| Year | UWTV abundance estimate | 95\% Confidence Interval | Landings in number | Total discards in number* | Removals in number | Harvest rate (by number)*** | Landings | Total discards* | Discard rate (by number) | Dead discard rate (by number) | Mean weight in landings | Mean <br> weight <br> in discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions |  |  |  |  | \% | tonnes |  | \% |  | grammes |  |
| 2002 | 1070 | 84 | 55 | 18 | 68 | 6.3 | 1154 | 192 | 24.5 | 19.6 | 21.2 | 10.8 |
| 2003 | 1246 | 187 | 44 | 18 | 58 | 4.6 | 933 | 183 | 29.3 | 23.7 | 21.2 | 10.0 |
| 2004 | 1410 | 108 | 29 | 11 | 38 | 2.7 | 525 | 112 | 28.2 | 22.9 | 18.1 | 9.9 |
| 2005 | 1092 | 62 | 42 | 20 | 57 | 5.2 | 778 | 182 | 31.7 | 25.9 | 18.4 | 9.2 |
| 2006 | 627 | 76 | n/a | n/a | 50 | 7.9 | 637 | n/a | n/a | n/a | n/a | n/a |
| 2007 | 920 | 62 | n/a | n/a | 57 | 6.2 | 913 | n/a | n/a | n/a | n/a | n/a |
| 2008 | 541 | 47 | 48 | 22 | 65 | 12.0 | 1057 | 248 | 31.4 | 25.6 | 21.94 | 11.23 |
| 2009 | 696 | 43 | 25 | 9 | 32 | 4.6 | 625 | 129 | 27.6 | 22.2 | 25.12 | 13.63 |
| 2010 | 879 | 47 | 37 | 15 | 49 | 5.6 | 939 | 224 | 29.0 | 23.4 | 25.16 | 14.70 |
| 2011 | 672 | 48 | 32 | 9 | 38 | 5.7 | 659 | 92 | 21.1 | 16.7 | 20.62 | 10.75 |
| 2012 | 468 | 52 | 61 | 8 | 67 | 14.4 | 1246 | 86 | 12.0 | 9.2 | 20.40 | 10.39 |
| 2013 | 441 | 65 | 60 | 12 | 69 | 15.7 | 1295 | 129 | 16.7 | 13.1 | 21.59 | 10.73 |
| 2014 | 383 | 57 | 34 | 5 | 38 | 9.8 | 766 | 48 | 12.9 | 10.0 | 22.62 | 9.56 |


| Year | UWTV abundance estimate | 95\% Confidence Interval | Landings in number | Total discards in number* | Removals in number | Harvest rate (by number)*** | Landings | Total discards* | Discard rate (by number) | Dead discard rate (by number) | Mean weight in landings | Mean weight in discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions |  |  |  |  | \% | tonnes |  | \% |  | grammes |  |
| 2015 | 556 | 71 | 18 | 2 | 19 | 3.4 | 370 | 15 | 8.4 | 6.4 | 20.91 | 9.13 |
| 2016 | 379 | 41 | 30 | 6 | 35 | 9.2 | 641 | 69 | 17.4 | 13.7 | 21.21 | 10.85 |
| 2017 | 404 | 41 | 13 | 4 | 16 | 4.0 | 295 | 38 | 21.4 | 16.9 | 22.23 | 10.46 |
| 2018 | 554 | 83 | 22 | 10 | 30 | 5.4 | 536 | 106 | 32.2 | 26.3 | 24.33 | 10.11 |
| 2019 | 493 | 66 | 8 | 2 | 9 | 1.9 | 167 | 21 | 21.7 | 17.2 | 22.00 | 9.94 |
| 2020 | 394 | 59 | 10 | 4 | 12 | 3.1 | 222 | 54 | 27.0 | 21.7 | 23.31 | 15.29 |
| 2021 | 331 | 31 | 23 | 7 | 28 | 8.5 | 498 | 88 | 23.6 | 18.9 | 21.88 | 12.48 |
| 2022 | 331* |  |  |  |  |  |  |  |  |  |  |  |

* 2022 abundance estimate used in the assessment was based on 2021 UWTV survey abundance estimate.


Figure 21.1.1. Nephrops in FU17 (Aran Grounds). Time-series of the number of Irish vessels reporting landings of Nephrops from FU17 with a >10 threshold.


Figure 21.1.2. Nephrops in FU17 (Aran Grounds). Combined box and kite plot of Irish vessel's power on the Aran Grounds by year. The blue line indicates the mean.


Figure 21.2.1. Nephrops in FU17 (Aran Grounds). Landings in tonnes by country.


Figure 21.2.2. Nephrops in FU17 (Aran Grounds). Effort data (kW days) for Irish directed Nephrops fleet.


Figure 21.2.3. Nephrops FU17 (Aran Grounds). Sampling levels for the Aran grounds.

## Length frequencies for catch (dotted) and landed(solid): <br> Nephrops in FU17



Figure 21.2.4. Nephrops FU17 Aran Grounds. Annual length composition of catches (dotted line) and landings (solid line) for females (left) and males (right) from 2008 (bottom) to 2020 (top). Annual mean length of catches (dotted vertical line) and landings (solid vertical line) are also shown. Minimum Landing Size ( 25 mm ) and 35 mm levels are also displayed with vertical lines.


Figure 21.2.5. Nephrops FU17 (Aran Grounds). Proportion of males by number in the catch (blue) and landings (black).


Figure 21.2.6. Nephrops FU17 (Aran Grounds). Mean weight in catch samples by sex showing cyclical trends.


Figure 21.2.7. Nephrops FU17 (Aran Grounds). Annual mean weight (g) estimates of landings (blue) and discards (black).


Figure 21.2.8. Nephrops in FU17 (Aran Grounds). Contour plots of the krigged density estimates for the Aran Ground UWTV surveys from 2002 (top left) to 2021 (bottom). No UWTV survey in 2022.


Figure 21.2.9. Nephrops FU17 Aran Grounds. Nephrops burrow estimates in FU17 Aran (blue), Galway Bay (green) and Slyne Head (red) grounds 2002-2021. 2022 UWTV was only carried out in Galway Bay.

## Stock size



Figure 21.2.10. Time-series of total abundance estimates for FU17.

## Fishing pressure



Figure 21.3.1. Nephrops FU17 Aran Grounds. Harvest Rate represented by red line (\% dead removed/UWTV abundance).

# 19 Norway lobster (Nephrops norvegicus) in divisions 7.a, 7.g and 7.j, Functional Unit 19 (Irish Sea, Celtic Sea, eastern part of southwest of Ireland) 

## Type of assessment in 2022

This stock was benchmarked in February 2014 and the assessment and provision of catch advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the benchmark WG (ICES, 2014) and set out in the stock annex. This stock assessment is available in the ICES Transparent Assessment Framework (TAF) here.

## ICES advice applicable to 2021

"ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, catches in 2021 that correspond to the $F$ ranges in the MAP are between 531 tonnes and 595 tonnes, assuming recent discard rates. The entire range is considered precautionary when applying the ICES advice rule.
To ensure that the stock in Functional Unit (FU) 19 is exploited sustainably, management should be implemented at the functional unit level."

## ICES advice applicable to 2022

"ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2018-2020, catches in 2022 that correspond to the F ranges in the MAP are between 337 and 378 tonnes.
To ensure that the stock in Functional Unit (FU) 19 is exploited sustainably, management should be implemented at the FU level."

### 19.1 General

## Stock description and management units

In FU19 Nephrops are caught on a large number of spatially discrete small inshore grounds and on some larger grounds further offshore and of these the 'Galley ground 4' and around Cork channels appear to be the most important (see Figure 19.1.1). The Nephrops stock (FU19) covers ICES rectangles ; 31-33 D9-E0; 31E1; 32E1-E2; 33E2-E3 within 7.a, 7.g, and 7.j. This stock is included as part of the TAC Area 7 Nephrops which includes the following stocks: Irish Sea East and West (FU14, FU15), Porcupine Bank (FU16), northwestern Irish Coast (FU18) and the Celtic Sea (FU20-22).
The map below shows FU19 assessment area (blue) and TAC area (red). There is no evidence that the individual functional units belong to the same stock. See Section 18 for details on Nephrops in Subarea 7 general section.


## Ecosystem aspects

This section is detailed in stock annex. There are no updates.

## Fishery description

A description of the fleet is given in the stock annex.
The time-series of numbers of vessels reporting landings greater than 10 t is updated in Figure 19.1.2. The numbers of vessels has been relatively stable from 1995 except since 2018, where there was a sharp decrease that has levelled. The time-series of vessel power is shown as a box and kite plot in Figure 19.1.3.

## Fishery in 2021

There has been a trend for Irish vessels ( $>18 \mathrm{~m}$ ) to switch to multi (quad) rig trawls. Provisional data suggest a $\sim 30 \%$ increase in Nephrops catch rates and a reduction in fish bycatch of $\sim 30 \%$ due to the lower headline height. The number of French vessels reporting landings in FU19, has decreased from 35 vessels in 2005 to five vessels in 2021.

## Information from stakeholders

None available.

### 19.2 Data

## InterCatch

All data were available in InterCatch and used for catch data only. French catch data provided directly by the national expert and not extracted from InterCatch.

## Landings

Landings data for FU19 are summarized in Table 19.2.1. Ireland, France and the UK report landings for FU19. Landings data for Ireland were revised back to 2008 which resulted in minor revisions in the order of 1 to $5 \%$ (stock annex). These revised data has been used in the assessment this year. The Republic of Ireland landings have fluctuated considerably throughout the timeseries, with a marked dip in 1994 (Table 19.2.1; Figure 19.2.1). The highest landings in the timeseries were observed in 2002-2004 (>1000 t). Landings in 2005 and 2006 have been below average for the series. In 2017 landings decreased by approximately $30 \%$ for the Irish fleet and were below the series average. This can be explained due to the poor weather conditions in quarter 1 which hampered fishing activities of smaller vessels and the larger vessels maximising effort in other FUs. There was a minor revision to 2018 landings for Ireland. Landing in 2021 were at a similar level to that reported in 2016. Landings by the French fleet have fluctuated with a declining trend throughout the time-series from the highest value in 1989 of 245 t to 1.1 t in 2020 . There was a minor revision to the $2019 \mathrm{UK}(\mathrm{E} \& \mathrm{~W})$ landings due to a code error (from 1.4 t to 1.1 t ). Landings from the UK are minor $<0.5 \mathrm{t}$ in 2020. This had a minimal effect on combined international data workup for that fishery year (Table 19.3.1.).

Total landings for years 2019 (value 249.1477 t ) and 2020 (value 248.9602 t ) are the same ( 249 t ) due to rounding.
WGCSE 2022 discovered a code error in year 2019 assessment where the international scaling was not carried out fully. This revision was presented to WGCSE 2022 and resulted in changes to numbers in landings, discards, removals, harvest rate and mean weights for year 2019. Table 19.3.1 is updated to reflect this change.

This data revision affects advice issued by WGCSE 2021 where the input year range is recent three year average (2018, 2019, 2020).

## Effort

In line with WGCSE 2015 recommendation effort is reported in KWdays and lpue reported in $\mathrm{KG} / \mathrm{kwdays}$ in the knowledge that the trend is likely to be a biased underestimate because it is not adjusted for efficiency or behavioural changes. The effort series is based on the same criteria for FU15, 16, 17, 22 and 20-21 ( $30 \%$ landings threshold) and will be contingent on the accuracy of landings data reported in logbooks.
Disaggregated effort and landings data are available for the Irish Nephrops directed fleet in FU19 from 1995-2021 for all vessels and vessels $>18$ metres total length. (Table 19.2.2; Figure 19.2.2). For vessels $>18$ effort (since early 2000s) has fluctuated with an overall decreasing trend in recent three years. This can be explained by fleet mobility where vessels target Nephrops in this area in periods of good emergence. For vessels <18 effort has decreased in 2017 to 2019 due to weather conditions.

## Sampling levels

Sampling levels, data aggregating and raising procedures were reviewed by WKCELT 2014, and are documented in the stock annex. The time-series of samples is shown in Figure 21.2.3 and Table 21.2.3. Sampling levels in 2021 were good and are comparable to recent levels.

## Commercial length-frequency distributions

Length-frequency data of the landings were collected on a regular basis from 2002 to 2020. Spatial and temporal coverage is problematic with landings from FU19 coming from several discrete grounds (see stock annex.) The sampling intensity and coverage has varied over the time-series (see stock annex). Since 2008 sampling has been good although the majority of the samples come from Bantry Bay recently. Also sampling of the discards is quite sparse over the time-series and are difficult to obtain due to the spatial coverage of the grounds. The catch samples from 2008 to 2021 were split using the discard selection ogive agreed at the benchmark. The length-weight regression parameters given in the stock annex are used to calculate sampled weights and appropriate quarterly raising factors. The length distributions are shown in Figure 19.2.4. The mean size has remained relatively stable and the trend in mean size is stable in recent years.

## Sex ratio

The sex ratio in the landings is male biased in most years but there is a trend towards increased percentage of females in the landings (Figure 19.2.5). The proportion of females was higher in 2013 and this was confirmed by the industry.

## Mean weight explorations

Explorations of the mean weight in the catch samples by sex shows a strong cyclical pattern in the females for all grounds combined (Figure 19.2.6). This corresponds with the emergence of mature females from the burrows to mate in summer. These data also show an increase in mean weights for males in 2016. The annual mean weight estimate for landings and discards is shown in Figure 19.2.7. The landings mean weight estimates increased in 2019 and then show a decrease in 2020.

## Discarding

Sampling of the discards has quite sparse over the time-series and are difficult to obtain due to the spatial coverage of the grounds (see stock annex). Since 2002 discard rates have been estimated using unsorted catch and discards sampling (as described in the stock annex). WKCELT 2014 examined the available discard data observations for FU19. An average discard selection ogive using data from Bantry Bay in years 2008 and 2013 was generated and deemed appropriate given the variable sampling intensity and coverage. The catch data from 2008 were then revised and split into landings and discards. Catch data sampling for years previous to 2008 was not revised as was considered to be not of good enough quality. The catch data were split using this selection ogive for the time series to date.

Discard rates range between $25-86 \%$ of total catch by weight and $40-80 \%$ of total catch by number (Table 19.2.4). These high discard rates are very high compared with other FUs. This is because the fleet is mainly smaller inshore vessels with limited space for extra crew. On-board "tailing" of the smaller Nephrops is not usually practised and the bigger Nephrops are picked from catches. There is no information on discard survival rate in this fishery but a $25 \%$ discard survival rate is assumed in line with other Nephrops stocks in the Celtic Sea.

Gear selectivity trials by Bord Iascaigh Mhara (BIM, 2017) reported a $64 \%$ survivor rate for Nephrops caught in a trawl with a SELTRA selectivity device in the outer Galway Bay area.

Table 19.3.1 gives weights, numbers and mean weights of the landings and discard raised internationally according to the stock annex.

## Abundance indices from UWTV surveys

The methods used during the survey were similar to those employed for UWTV surveys [U5917] of Nephrops stocks around Ireland and elsewhere are documented by WKNEPHTV (ICES, 2007), WKNEPHBID (ICES, 2008), SGNEPS (ICES, 2009; 2010; 2012), WGNEPS (ICES, 2013; 2014; 2015; 2016a; 2017; 2018a, 2020, 2021, 2022), WKNEPS (ICES, 2016b; 2018b), Leocádio, A., et al, 2018 and Dobby H., et al, 2021. SGNEPS 2012 (ICES, 2012) recommended that a CV (or relative standard error) of $<20 \%$ as an acceptable precision level for UWTV survey estimates of abundance. Given the scale of the area and the number of distinct patches it is unrealistic to expect sufficient stations $(\sim 10)$ in each individual patch to estimate densities separately. The random stratified approach may cause problems in years where the planned survey coverage is not achieved. WKCELT 2014 concluded that WGCSE or WGNEPS should make recommendations on the most appropriate fill in procedure to be adopted in these cases.

The spatial extent of the Nephrops grounds in FU19 has been re-defined by WKCELT 2014 and the abundance estimates are calculated using these areas. The redefinition of the polygons in FU19 resulted in $\sim 16 \%$ increase in overall area from $1653 \mathrm{~km}^{2}$ to $1973 \mathrm{~km}^{2}$ (see stock annex). The discrete grounds have been named as: Bantry Bay, Galley Ground 1-4, Cork Channels and Helvick 1-2 and are shown in Figure 19.1.1. In terms of area the Galley Grounds (1-4) account for $61 \%$ of the total grounds in FU19 and Galley Ground 4 is the largest of these representing $47 \%$ of the total area (Table 19.2.5). Helvick patches 2 and 3 were also amalgamated and renamed Helvick 2 based on the information from the VMS data.

From 2011 to 2022 an average of 42 stations have been completed annually. The survey design is based on randomly picked stations from the ground polygons and the sampling effort on each ground was determined by relative area.

All grounds except Galley Ground 4 in 2011 and Galley Ground 1 in 2012 were covered by the TV survey. Since 2015 a new patch Kenmare Bay was surveyed.
Detailed summary statistics for the various Nephrops patches in FU19 over the time-series are presented in Table 21.2.6. The mean density varies across the different patches, but there is some consistency to the estimates over time. In 2022 all discrete grounds were covered by the TV survey (Doyle et al., 2022).

The 2022 mean density estimates vary between patches from the lowest value 0.04 (no. $/ \mathrm{m}^{2}$ ) observed at Kenmare Bay to the highest observed at 0.39 (no. $/ \mathrm{m}^{2}$ ) at Galley ground 2 (Table 19.2.6, Figure 19.2.8). The overall mean density for FU19 in 2022 is 0.13 (no. $/ \mathrm{m}^{2}$ ) which is the lowest observed in the time-series (Table 19.2.7).

Figure 19.2.9 and Table 19.2.7 shows the total abundance estimate for FU19 with the WKMSYRef4 proposed MSY Btrigger (ICES, 2016XX, ICESYY). The 2022 abundance estimate was $4 \%$ lower than in 2021 and at 259 million is below the MSY B trigger ( 430 million) with a RSE of $14 \%$ which is below the $20 \%$ limit recommended by SGNEPs (2012).

## Information from Irish Groundfish survey

Length-frequency data of the Nephrops catches on the Irish groundfish survey-Q4: IGFS-WIBTSQ4 [G7212] from 2003-2021 are available (Stokes et al., 2014; ICES, 2015). These data were investigated for trends in indicators such as possible recruitment signals (Figure 19.2.10). The mean size of males and females in from the survey was fairly stable over time at 33 mm for males and 25 mm for females.

### 19.3 Assessment

## Comparison with previous assessments

The WGCSE 2019 carried out an UWTV based assessment for this stock. The methods used were very much in line with WKNEPH (ICES, 2009) and the approach taken for other Nephrops stocks in 6 and 7 by WGCSE. This approach was benchmarked at WKCELT 2014 (ICES, 2014).

## State of the stock

UWTV abundance estimates suggest that the stock size has fluctuated with a declining trend in the recent five years. The 2022 estimate is the lowest observed and is below the MSY $\mathrm{B}_{\text {trigger. }}$. The 2022 abundance remains below the average of the series (geomean: [2011-2022]: 401 million).

Table 19.3.1 summarizes recent abundance estimates, harvest rates for the stock along with other stock parameters. Harvest rate is calculated as (landings + dead discards)/(abundance estimate).

Table 19.3.1.and Figure 19.3 .1 summarize recent harvest ratios which have been below the FMSY proxy for the last three years.

### 19.4 Catch scenario table

Catch scenario table inputs and historical estimates of mean weight in landings and harvest ratios are presented in Table 19.3.1 and summarised below.

The basis for the catch options:

| Variable | Value | Notes |
| :--- | :---: | :--- |
| Stock abundance (2023) | 259 | Numbers of individuals (millions); UWTV survey 2022 |
| Mean weight in projected landings | 27.2 | Average 2019-2021in grammes |
| Mean weight in projected discards | 13.5 | Average 2019-2021 in grammes |
| Projected discards | 48.6 | Proportion by number; average 2019-2021 |
| Discards survival | 25 | Proportion by number |
| Projected dead discards | 41.5 | Proportion by number; average 2019-2021 |

The average in the recent three years is used to calculate the mean weight for landings and discards. The discard rates and proportions for the last three years are used to account for recent on-board retention practices (this is also according to the stock annex).

A prediction of landings for the FU19 using the approach agreed procedure proposed at WKNEPH 2009 and outlined in the stock annex will be made on the basis of the 2022 UWTV survey. This will be presented in October 2022 for the provision of advice.

### 19.5 Reference points

WKMSYRef4 updated the Fmš reference points for FU19 (ICES, 2016XX; 2016YY) on the basis of an average of estimated FmSy proxy harvest rates over a period of years, this corresponds more
closely to the methodology for finfish. The updated harvest rate calculated at $9.3 \%$ is expected to deliver high long-term yield with a low probability of recruitment overfishing. This is close to the harvest rate of $8.1 \%$ calculated by WKCELT (ICES, 2014)

This stock previously did not have MSY Btrigger specified, the time-series and range of indicator biomass is also limited such that direct use of Bloss is considered too close to equilibrium biomass. The workshop proposed to use the $5 \%$ interval on the probability distribution of indicator biomass assuming a normal distribution, which is analogous to the $5 \%$ on Bmsy proposed for finfish stocks assuming these Nephrops FU have been exploited at a rate close to near HRmsy. The MSY $B_{\text {trigger }}$ for FU 19 is 434 million individuals rounded to 430 million.

These reference points shown in text table below should remain under review by WGCSE should improved data become available.

| Stock code | MSY Flower* | FMSY* | MSY Fupper* with AR | MSY Btrigger | MSY Fupper* with no AR |
| :--- | :---: | :---: | :---: | :---: | :---: |
| nep-19 | $8.3 \%$ | $9.3 \%$ | $9.3 \%$ | $430^{* * *}$ | $9.3 \%$ |

* Harvest rate (HR).
*** Abundance in millions.


### 19.6 Management strategies

No specific management plan exists for this stock.
The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to Norway lobster (Nephrops norvegicus) by functional unit in ICES subarea 7 and also demersal stocks.

### 19.7 Quality of assessment and forecast

Biological sampling for this stock is improving given the spatial distribution of the Nephrops mud patches. A number of other biological parameters such as mean weights and length distributions have also been revised. The revisions were made as part of the benchmark process and have improved the quality of the assessment.

In the provision of catch options based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. For FU19 deterministic estimates of the mean weight in the landings and discard rates for 2019-2021 are used although there is some variability of these over time.
From 2016, fisheries catching Nephrops in Subarea 7 are covered by the EU landings obligation (EU, 2015). Creel fisheries are exempted from the landings obligation, with a de minimis exemption consisting of a $5 \%$ discard rate by weight for the trawl fishery in 2019 (reduced from $6 \%$ in 2018 and $7 \%$ in both 2016 and 2017).

Irish discard survival experiments indicate that the trawl discard survival may be around $64 \%$ (BIM, 2017). As a result, an exemption from the landings obligation based on high survivability has been granted by the European Commission. The average discard rate by weight for FU19 over the last three years is $32 \%$. Catch advice and scenarios are provided this year on the assumption that discarding is assumed to continue at the recent average.

There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008; SGNEPS 2009; WGNEPS 2014). Ultimately there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate, although no more precise WKNEPH (ICES, 2009). Different densities are apparent on the various different grounds within this FU. For the 2022 survey the number of observations on each individual patch is relatively low making the relative standard error (RSE) estimates not that relevant. Aggregating all areas together gives a mean burrow density of 0.13 with a RSE of around $14 \%$ which is below the $20 \%$ threshold recommended by SGNEPS (ICES, 2012). The cumulative bias estimates for FU19 are largely based on expert opinion. The precision of these bias corrections cannot yet be characterized, but is likely to be lower than that observed in the survey.

Landings data are adjusted to take into account landings that have been misreported from FU16 since 2011. This adjustment is thought to be reasonably accurate (See Section 19).

### 19.8 Recommendations for next benchmark

This stock was benchmarked by ICES in February 2014 (ICES, 2014). WGCSE will keep the stock under close review and recommend future benchmark as required.

### 19.9 Management considerations

The trends from the fishery (landings, effort, mean size, etc.) appear to show a decline. The UWTV abundance and mean density estimates vary between the discrete patches and population dynamics between these are not fully understood. The 2022 survey result is the lowest observed in the time-series.

In recent years several newer vessels specializing in Nephrops fishing have participated in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates. Since the introduction of effort management associated with the cod long-term plan (EC 1342/2008) there have been concerns that effort will be displaced towards FU19 and other Nephrops grounds where effort control has not been put in place.
Nephrops fisheries in this area are fairly mixed also catching megrim, anglerfish and other demersal species. There are also some catches of hake, and in the offshore parts of the area. The Nephrops grounds in FU19 coincide with an important nursery area for juvenile hake and anglerfish among other species (ICES, 2009).

### 19.10 References

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Table 19.2.1. Nephrops in FU19 (SW and SE Ireland). Landings in tonnes by country. The figures in the table are rounded. Calculations were done with unrounded inputs and computed values may not match exactly when calculated using the rounded figures in the table.

| Year | FU 19 |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | France | Rep. of Ireland | UK | Total |
| 1989 | 245 | 652 | 2 | 899 |
| 1990 | 181 | 569 | 4 | 754 |
| 1991 | 212 | 860 | 5 | 1077 |
| 1992 | 233 | 640 | 15 | 888 |
| 1993 | 229 | 672 | 4 | 905 |
| 1994 | 216 | 153 | 21 | 390 |
| 1995 | 175 | 507 | 12 | 694 |
| 1996 | 145 | 736 | 7 | 888 |
| 1997 | 93 | 656 | 7 | 756 |
| 1998 | 92 | 733 | 2 | 827 |
| 1999 | 77 | 499 | 3 | 579 |
| 2000 | 144 | 541 | 11 | 696 |
| 2001 | 111 | 702 | 2 | 815 |
| 2002 | 188 | 1130 | 0 | 1318 |
| 2003 | 165 | 1075 | 0 | 1240 |
| 2004 | 76 | 997 | 1 | 1074 |
| 2005 | 62 | 648 | 2 | 712 |
| 2006 | 65 | 675 | 1 | 741 |
| 2007 | 63 | 894 | 0 | 957 |
| 2008 | 46 | 790 | 15 | 851 |
| 2009 | 55 | 798 | 15 | 868 |
| 2010 | 14 | 660 | 13 | 687 |
| 2011 | 23 | 619 | 1 | 643 |


| Year | FU 19 |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| 2012 | France | Rep. of Ireland | Total |  |
| 2013 | 11 | 837 | 1 | 849 |
| 2014 | 4 | 783 | 3 | 794 |
| 2015 | 5 | 502 | 3 | 468 |
| 2016 | 4 | 483 | 4 | 507 |
| 2017 | 4 | 229 | 247 | 1 |

Table 19.2.2. Nephrops in FU19 (SW and SE Ireland). Irish Nephrops directed effort (Kw Days) and landings. Irish Fleet Nephrops trawlers (>30\% landings weight)

| Year |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | All Vessels |  | Vessels >18 m |  |
|  | kW days ('000) | Landings Tonnes | kW days ('000) | Landings Tonnes |
| 1995 | 222.0 | 380 | 80.7 | 121 |
| 1996 | 178.6 | 355 | 55.6 | 86 |
| 1997 | 161.0 | 306 | 53.9 | 101 |
| 1998 | 329.6 | 498 | 144.6 | 189 |
| 1999 | 182.9 | 236 | 42.3 | 47 |
| 2000 | 142.0 | 217 | 56.2 | 86 |
| 2001 | 193.3 | 397 | 89.1 | 139 |
| 2002 | 506.7 | 883 | 323.7 | 446 |
| 2003 | 555.9 | 693 | 318.8 | 364 |
| 2004 | 488.1 | 558 | 303.0 | 311 |
| 2005 | 405.0 | 471 | 220.6 | 219 |
| 2006 | 424.2 | 478 | 208.8 | 186 |
| 2007 | 558.8 | 713 | 287.4 | 262 |
| 2008 | 534.1 | 643 | 288.1 | 319 |
| 2009 | 472.0 | 613 | 224.5 | 243 |
| 2010 | 382.2 | 494 | 103.7 | 114 |
| 2011 | 337.3 | 449 | 142.9 | 167 |
| 2012 | 355.5 | 541 | 91.9 | 126 |
| 2013 | 336.1 | 571 | 88.6 | 133 |
| 2014 | 213.6 | 332 | 52.1 | 74 |
| 2015 | 244.6 | 393 | 85.5 | 118 |
| 2016 | 287.3 | 558 | 111.2 | 233 |
| 2017 | 118.2 | 425 | 111.4 | 179 |
| 2018 | 71.6 | 107.1 | 24.1 | 29.9 |
| 2019 | 91.4 | 145.9 | 31.6 | 37.5 |


| Year |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
|  | All Vessels | Vessels $>18 \mathrm{~m}$ |  |  |
|  | kW days ('000) | Landings Tonnes | kW days ('000) | Landings Tonnes |
| 2020 | 72.3 | 133.4 | 12.7 | 19.1 |
| 2021 | 125 | 261 | 29 | 47 |

Table 19.2.3. Nephrops in FU19 (SW and SE Ireland). Irish Sampling levels.

| Year | Quarter | Number of samples |  |  | Numbers Measured |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Landings | Catch | Discards | Landings |
| 2008 | 1 | 3 | 0 | 0 | 1502 | 0 | 0 |
| 2008 | 2 | 6 | 0 | 0 | 3521 | 0 | 0 |
| 2008 | 3 | 6 | 0 | 0 | 6412 | 0 | 0 |
| 2008 | 4 | 3 | 0 | 0 | 876 | 0 | 0 |
| 2009 | 1 | 3 | 0 | 0 | 1347 | 0 | 0 |
| 2009 | 2 | 6 | 0 | 0 | 3369 | 0 | 0 |
| 2009 | 3 | 2 | 0 | 0 | 1003 | 0 | 0 |
| 2009 | 4 | 5 | 0 | 0 | 1882 | 0 | 0 |
| 2010 | 1 | 2 | 0 | 0 | 840 | 0 | 0 |
| 2010 | 2 | 7 | 0 | 0 | 2989 | 0 | 0 |
| 2010 | 3 | 4 | 0 | 0 | 1457 | 0 | 0 |
| 2010 | 4 | 6 | 0 | 0 | 2376 | 0 | 0 |
| 2011 | 1 | 3 | 0 | 0 | 1493 | 0 | 0 |
| 2011 | 2 | 5 | 0 | 0 | 2747 | 0 | 0 |
| 2011 | 3 | 2 | 0 | 0 | 938 | 0 | 0 |
| 2011 | 4 | 5 | 0 | 0 | 2686 | 0 | 0 |
| 2012 | 1 | 6 | 0 | 0 | 2053 | 0 | 0 |
| 2012 | 2 | 7 | 0 | 0 | 3956 | 0 | 0 |
| 2012 | 3 | 4 | 0 | 0 | 1980 | 0 | 0 |
| 2012 | 4 | 4 | 0 | 0 | 1969 | 0 | 0 |


| Year | Quarter | Number of samples |  |  | Numbers Measured |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Landings | Catch | Discards | Landings |
| 2013 | 1 | 3 | 0 | 0 | 1857 | 0 | 0 |
| 2013 | 2 | 8 | 5 | 0 | 4117 | 2059 | 0 |
| 2013 | 2 | 3 | 3 | 0 | 1177 | 1250 | 0 |
| 2013 | 4 | 3 | 3 | 0 | 1472 | 1276 | 0 |
| 2014 | 1 | 3 | 2 | 0 | 1137 | 941 | 0 |
| 2014 | 2 | 7 | 7 | 0 | 3331 | 2319 | 0 |
| 2014 | 3 | 3 | 2 | 0 | 1344 | 682 | 0 |
| 2014 | 4 | 10 | 8 | 0 | 3455 | 2200 | 0 |
| 2015 | 1 | 1 | 1 | 0 | 417 | 310 | 0 |
| 2015 | 2 | 3 | 3 | 0 | 1417 | 1267 | 0 |
| 2015 | 3 | 2 | 2 | 1 | 856 | 648 | 321 |
| 2015 | 4 | 3 | 2 | 0 | 1250 | 774 | 0 |
| 2016 | 1 | 3 | 3 | 0 | 1500 | 1631 | 0 |
| 2016 | 2 | 6 | 5 | 0 | 2310 | 1760 | 0 |
| 2016 | 3 | 9 | 7 | 0 | 3328 | 2448 | 0 |
| 2016 | 4 | 5 | 5 | 0 | 1,923 | 1521 | 0 |

Table 19.2.3. Continued.

| Year | Quarter | Number of samples |  |  | Numbers Measured |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Landings | Catch | Discards | Landings |
| 2017 | 1 | 4 | 4 | 0 | 1860 | 1283 | 0 |
| 2017 | 2 | 3 | 3 | 0 | 1572 | 1281 | 0 |
| 2017 | 3 | 2 | 2 | 0 | 998 | 943 | 0 |
| 2017 | 4 | 4 | 2 | 0 | 1200 | 785 | 0 |
| 2018 | 1 | 1 | 1 | 0 | 304 | 380 | 0 |
| 2018 | 2 | 7 | 7 | 0 | 3579 | 3230 | 0 |
| 2018 | 3 | 1 | 1 | 0 | 255 | 275 | 0 |
| 2018 | 4 | 1 | 1 | 0 | 370 | 404 | 0 |
| 2019 | 1 | 4 | 5 | 0 | 1630 | 2222 | 0 |


| Year | Quarter | Number of samples |  |  | Numbers Measured |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Landings | Catch | Discards | Landings |
| 2019 | 2 | 3 | 3 | 0 | 1275 | 1398 | 0 |
| 2019 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 4 | 4 | 4 | 0 | 1810 | 1798 | 0 |
| 2020 | 1 | 2 | 2 | 0 | 728 | 702 | 0 |
| 2020 | 2 | 7 | 7 | 0 | 3095 | 2855 | 0 |
| 2020 | 3 | 1 | 1 | 0 | 489 | 404 | 0 |
| 2020 | 4 | 3 | 4 | 0 | 1671 | 1900 | 0 |
| 2021 | 1 | 2 | 2 | 0 | 842 | 782 | 0 |
| 2021 | 2 | 5 | 5 | 0 | 2530 | 2484 | 0 |
| 2021 | 3 | 3 | 3 | 0 | 1497 | 1326 | 0 |
| 2021 | 4 | 4 | 4 | 0 | 2363 | 2415 | 0 |

Table 19.2.4. Nephrops in FU19 (SW and SE Ireland). Landings and estimated discards by weight.

|  | Female |  | Male |  | Both sexes |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings (t) | Discards (t) | Landings (t) | Discards (t) | \% Discard |
| 2008 | 99 | 29 | 691 | 69 | 11.0 |
| 2009 | 117 | 106 | 681 | 141 | 23.7 |
| 2010 | 138 | 98 | 522 | 148 | 27.2 |
| 2011 | 169 | 155 | 450 | 250 | 38.9 |
| 2012 | 190 | 202 | 647 | 265 | 35.8 |
| 2013 | 259 | 210 | 525 | 220 | 35.4 |
| 2014 | 106 | 71 | 353 | 87 | 25.6 |
| 2015 | 79 | 64 | 423 | 101 | 24.8 |
| 2016 | 154 | 91 | 429 | 100 | 24.7 |
| 2017 | 133 | 58 | 280 | 79 | 24.9 |
| 2018 | 71 | 27 | 157 | 40 | 22.9 |
| 2019 | 66 | 48 | 181 | 63 | 31.1 |
| 2020 | 40 | 46 | 207 | 89 | 35.3 |
| 2021 | 83 | 63 | 331 | 109 | 29.4 |

Table 19.2.5. Nephrops in FU19 (SW and SE Ireland). Landings and estimated discards by number.

|  | Female Numbers '000s |  | Male Numbers '000s |  | Both sexes |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings | Discards | Landings | Discards | \% Discard |
| 2008 | 3,893 | 1,781 | 19,516 | 3,255 | 17.7 |
| 2009 | 5,819 | 8,250 | 20,324 | 8,793 | 39.5 |
| 2010 | 6,276 | 8,147 | 16,001 | 10,117 | 45.1 |
| 2011 | 7,295 | 12,895 | 16,900 | 18,192 | 55.7 |
| 2012 | 9,266 | 17,635 | 22,540 | 19,108 | 53.6 |
| 2013 | 11,680 | 18,945 | 17,399 | 17,034 | 55.3 |
| 2014 | 4,862 | 5,647 | 11,183 | 5,572 | 41.1 |
| 2015 | 3,706 | 5,255 | 13,111 | 6,462 | 41.1 |
| 2016 | 6,877 | 6,761 | 12,610 | 6,668 | 40.8 |
| 2017 | 5,295 | 4,400 | 9,022 | 5,044 | 39.7 |
| 2018 | 2,908 | 1,866 | 5,197 | 2,454 | 34.8 |
| 2019 | 2,970 | 3,909 | 6,023 | 4,474 | 48.2 |
| 2020 | 2,006 | 3,971 | 7,595 | 6,026 | 51.0 |
| 2021 | 3,701 | 5,133 | 10,817 | 7,481 | 46.5 |

Table 19.2.6. Nephrops in FU19 (SW and SE Ireland). Area ( $\mathrm{Km}^{2}$ ) of discrete patches and percentage contribution to overall area.

| Ground | Area $\left(\mathrm{Km}^{2}\right)$ | \% Contribution |
| :--- | :---: | :---: |
| Bantry | 121.5 | $6 \%$ |
| Cork Channels | 562.0 | $28 \%$ |
| Galley Grounds 1 | 60.9 | $3 \%$ |
| Galley Grounds 2 | 76.7 | $4 \%$ |
| Galley Grounds 3 | 133.9 | $4 \%$ |
| Galley Grounds 4 | 925.1 | $2 \%$ |
| Helvick 1 | 33.1 | $3 \%$ |
| Helvick 2 | 59.5 | $49 \%$ |
| Total | 1972.8 | 3 |

Table 19.2.7. Nephrops in FU19 (SW and SE Ireland). Detailed summary statistics for the various Nephrops patches in FU19 over the time-series. ( $\mathbf{N}=$ number of stations, Mean Density (burrow/m²) is adjusted for the bias correction factor in Table 3, sd, se and ci are the standard deviation, standard error and $95 \%$ confidence intervals on the mean density).

| Year | Ground | N | Mean Density | sd | se | ci |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | Galley Grounds 4 | 6 | 0.21 | 0.18 | 0.08 | 0.19 |
| 2011 | Bantry | 5 | 0.33 | 0.23 | 0.1 | 0.28 |
| 2011 | Cork Channels | 12 | 0.35 | 0.32 | 0.09 | 0.2 |
| 2011 | Galley Grounds 1 | 3 | 0.52 | 0.41 | 0.24 | 1.02 |
| 2011 | Galley Grounds 2 | 3 | 0.59 | 0.43 | 0.25 | 1.07 |
| 2011 | Galley Grounds 3 | 4 | 0.58 | 0.22 | 0.11 | 0.35 |
| 2011 | Helvick 1 | 3 | 0.6 | 0.01 | 0.01 | 0.04 |
| 2011 | Helvick 2 | 5 | 0.12 | 0.21 | 0.09 | 0.26 |
| 2012 | Bantry | 1 | 0.2 | NA | NA | NA |
| 2012 | Cork Channels | 9 | 0.27 | 0.17 | 0.06 | 0.13 |
| 2012 | Galley Grounds 2 | 4 | 0.59 | 0.12 | 0.06 | 0.19 |
| 2012 | Galley Grounds 3 | 1 | 0.51 | NA | NA | NA |
| 2012 | Galley Grounds 4 | 16 | 0.39 | 0.16 | 0.04 | 0.09 |
| 2012 | Helvick 1 | 3 | 0.33 | 0.13 | 0.08 | 0.33 |
| 2012 | Helvick 2 | 6 | 0.33 | 0.41 | 0.17 | 0.43 |
| 2013 | Bantry | 4 | 0.38 | 0.2 | 0.1 | 0.31 |
| 2013 | Cork Channels | 11 | 0.12 | 0.1 | 0.03 | 0.07 |
| 2013 | Galley Grounds 1 | 2 | 0.23 | 0.18 | 0.13 | 1.59 |
| 2013 | Galley Grounds 2 | 3 | 0.48 | 0.44 | 0.25 | 1.09 |
| 2013 | Galley Grounds 3 | 4 | 0.59 | 0.24 | 0.12 | 0.38 |
| 2013 | Galley Grounds 4 | 13 | 0.19 | 0.27 | 0.07 | 0.16 |
| 2013 | Helvick 1 | 1 | 0.09 | NA | NA | NA |
| 2013 | Helvick 2 | 2 | 0.06 | 0.05 | 0.04 | 0.48 |
| 2014 | Bantry | 4 | 0.25 | 0.05 | 0.03 | 0.09 |
| 2014 | Cork Channels | 10 | 0.1 | 0.06 | 0.02 | 0.04 |
| 2014 | Galley Grounds 1 | 2 | 0.61 | 0.41 | 0.29 | 3.69 |
| 2014 | Galley Grounds 2 | 2 | 0.82 | 0.14 | 0.1 | 1.23 |
| 2014 | Galley Grounds 3 | 4 | 0.66 | 0.23 | 0.12 | 0.37 |


| Year | Ground | N | Mean Density | sd | se | ci |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | Galley Grounds 4 | 14 | 0.29 | 0.29 | 0.08 | 0.17 |
| 2014 | Helvick 1 | 2 | 0.67 | 0.28 | 0.2 | 2.53 |
| 2014 | Helvick 2 | 2 | 0.03 | 0.04 | 0.03 | 0.39 |
| 2015 | Bantry | 2 | 0.32 | 0.11 | 0.08 | 1.02 |
| 2015 | Cork Channels | 10 | 0.08 | 0.11 | 0.03 | 0.08 |
| 2015 | Galley Grounds 1 | 2 | 0.32 | 0.46 | 0.32 | 4.12 |
| 2015 | Galley Grounds 2 | 2 | 0.53 | 0.08 | 0.06 | 0.74 |
| 2015 | Galley Grounds 3 | 4 | 0.40 | 0.14 | 0.07 | 0.23 |
| 2015 | Galley Grounds 4 | 14 | 0.27 | 0.19 | 0.05 | 0.11 |
| 2015 | Helvick 1 | 2 | 0.30 | 0.23 | 0.16 | 2.08 |
| 2015 | Helvick 2 | 2 | 0.09 | 0.09 | 0.06 | 0.79 |
| 2015 | Kenmare Bay | 1 | 0.30 | NA | NA | NA |
| 2016 | Bantry | 4 | 0.20 | 0.07 | 0.04 | 0.12 |
| 2016 | Cork Channels | 10 | 0.21 | 0.11 | 0.03 | 0.08 |
| 2016 | Galley Grounds 1 | 2 | 0.03 | 0.01 | 0.01 | 0.08 |
| 2016 | Galley Grounds 2 | 2 | 0.53 | 0.12 | 0.09 | 1.11 |
| 2016 | Galley Grounds 3 | 4 | 0.16 | 0.12 | 0.06 | 0.19 |
| 2016 | Galley Grounds 4 | 14 | 0.17 | 0.20 | 0.05 | 0.12 |
| 2016 | Helvick 1 | 2 | 0.38 | 0.08 | 0.06 | 0.70 |
| 2016 | Helvick 2 | 2 | 0.07 | 0.09 | 0.06 | 0.81 |
| 2016 | Kenmare Bay | 2 | 0.24 | 0.15 | 0.11 | 1.33 |
| 2017 | Bantry | 3 | 0.29 | 0.15 | 0.09 | 0.37 |
| 2017 | Cork Channels | 10 | 0.25 | 0.20 | 0.06 | 0.14 |
| 2017 | Galley Grounds 1 | 2 | 0.24 | 0.11 | 0.08 | 1.00 |
| 2017 | Galley Grounds 2 | 2 | 0.63 | 0.06 | 0.04 | 0.55 |
| 2017 | Galley Grounds 3 | 3 | 0.45 | 0.12 | 0.07 | 0.30 |
| 2017 | Galley Grounds 4 | 15 | 0.16 | 0.16 | 0.04 | 0.09 |
| 2017 | Helvick 1 | 2 | 0.46 | 0.07 | 0.05 | 0.66 |
| 2017 | Helvick 2 | 2 | 0.16 | 0.23 | 0.16 | 2.03 |


| Year | Ground | N | Mean Density | sd | se | ci |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | Kenmare Bay | 2 | 0.16 | 0.22 | 0.16 | 1.97 |
| 2018 | Bantry | 4 | 0.06 | 0.02 | 0.01 | 0.04 |
| 2018 | Cork Channels | 10 | 0.11 | 0.11 | 0.04 | 0.08 |
| 2018 | Galley Grounds 1 | 2 | 0.06 | 0.01 | 0.01 | 0.10 |
| 2018 | Galley Grounds 2 | 2 | 0.19 | 0.19 | 0.14 | 1.75 |
| 2018 | Galley Grounds 3 | 4 | 0.11 | 0.09 | 0.05 | 0.14 |
| 2018 | Galley Grounds 4 | 14 | 0.07 | 0.08 | 0.02 | 0.05 |
| 2018 | Helvick 1 | 2 | 0.11 | 0.10 | 0.07 | 0.92 |
| 2018 | Helvick 2 | 2 | 0.06 | 0.03 | 0.02 | 0.28 |
| 2018 | Kenmare Bay | 2 | 0.07 | 0.03 | 0.02 | 0.25 |
| 2019 | Bantry | 4 | 0.13 | 0.04 | 0.02 | 0.06 |
| 2019 | Cork Channels | 10 | 0.16 | 0.17 | 0.06 | 0.13 |
| 2019 | Galley Grounds 1 | 2 | 0.12 | 0.17 | 0.12 | 1.57 |
| 2019 | Galley Grounds 2 | 2 | 0.66 | 0.38 | 0.27 | 3.40 |
| 2019 | Galley Grounds 3 | 4 | 0.21 | 0.14 | 0.07 | 0.23 |
| 2019 | Galley Grounds 4 | 14 | 0.18 | 0.23 | 0.06 | 0.13 |
| 2019 | Helvick 1 | 2 | 0.34 | 0.27 | 0.19 | 2.46 |
| 2019 | Helvick 2 | 2 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | Kenmare Bay | 2 | 0.27 | 0.10 | 0.07 | 0.88 |
| 2020 | Bantry | 0.31 | 0.11 | 0.05 | 0.17 | 0.31 |
| 2020 | Cork Channels | 0.13 | 0.20 | 0.06 | 0.14 | 0.13 |
| 2020 | Galley Grounds 1 | 0.13 | 0.10 | 0.07 | 0.87 | 0.13 |
| 2020 | Galley Grounds 2 | 0.43 | 0.24 | 0.17 | 2.14 | 0.43 |
| 2020 | Galley Grounds 3 | 0.20 | 0.15 | 0.08 | 0.24 | 0.20 |
| 2020 | Galley Grounds 4 | 0.10 | 0.10 | 0.03 | 0.06 | 0.10 |
| 2020 | Helvick 1 | 0.24 | 0.05 | 0.04 | 0.48 | 0.24 |
| 2020 | Helvick 2 | 0.06 | 0.08 | 0.06 | 0.73 | 0.06 |
| 2020 | Kenmare Bay | 0.18 | 0.12 | 0.09 | 1.11 | 0.18 |
| 2021 | Bantry | 4 | 0.09 | 0.03 | 0.01 | 0.04 |

Table 19.2.7. Continued.

| Year | Ground | N | Mean Density | sd | se | ci |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2021 | Cork Channels | 10 | 0.20 | 0.19 | 0.06 | 0.14 |
| 2021 | Galley Grounds 1 | 2 | 0.08 | 0.06 | 0.04 | 0.54 |
| 2021 | Galley Grounds 2 | 2 | 0.31 | 0.10 | 0.07 | 0.87 |
| 2021 | Galley Grounds 3 | 4 | 0.22 | 0.13 | 0.06 | 0.20 |
| 2021 | Galley Grounds 4 | 14 | 0.09 | 0.07 | 0.02 | 0.04 |
| 2021 | Helvick 1 | 2 | 0.09 | 0.08 | 0.05 | 0.69 |
| 2021 | Helvick 2 | 2 | 0.08 | 0.05 | 0.04 | 0.48 |
| 2021 | Kenmare Bay | 2 | 0.05 | 0.03 | 0.02 | 0.30 |
| 2022 | Bantry | 4 | 0.08 | 0.06 | 0.03 | 0.10 |
| 2022 | Cork Channels | 10 | 0.10 | 0.13 | 0.04 | 0.09 |
| 2022 | Galley Grounds 1 | 2 | 0.06 | 0.01 | 0.01 | 0.13 |
| 2022 | Galley Grounds 2 | 2 | 0.39 | 0.26 | 0.19 | 2.35 |
| 2022 | Galley Grounds 3 | 4 | 0.17 | 0.05 | 0.03 | 0.08 |
| 2022 | Galley Grounds 4 | 14 | 0.15 | 0.11 | 0.03 | 0.07 |
| 2022 | Helvick 1 | 2 | 0.14 | 0.00 | 0.00 | 0.02 |
| 2022 | Helvick 2 | 2 | 0.08 | 0.08 | 0.06 | 0.74 |
| 2022 | Kenmare Bay | 2 | 0.04 | 0.05 | 0.04 | 0.46 |

Table 19.2.7. Nephrops in FU19 (SW and SE Ireland). Summary statistics for FU19 combined over the time-series.

| Year | Number of stations | Mean Density adjusted (burrow $/ \mathrm{m}^{2}$ ) | Standard Deviation | Raised abundance estimate adjusted (million burrows) | Upper 95\%CI on Abundance | Lower 95\%CI on Abundance | CVs <br> (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 6 | 0.21 | 0.18 | 408 | 789 | 26 | 36 |
| 2007* |  |  |  |  |  |  |  |
| 2008* |  |  |  |  |  |  |  |
| 2009* |  |  |  |  |  |  |  |
| 2010* |  |  |  |  |  |  |  |
| 2011 | 35 | 0.34 | 0.26 | 665 | 836 | 494 | 13 |
| 2012 | 40 | 0.3 | 0.18 | 594 | 705 | 484 | 9 |
| 2013 | 40 | 0.25 | 0.26 | 487 | 648 | 326 | 17 |
| 2014 | 40 | 0.32 | 0.31 | 636 | 823 | 448 | 15 |
| 2015 | 39 | 0.24 | 0.2 | 482 | 608 | 356 | 13 |
| 2016 | 42 | 0.2 | 0.17 | 399 | 498 | 299 | 13 |
| 2017 | 41 | 0.25 | 0.20 | 499 | 619 | 379 | 12 |
| 2018 | 42 | 0.09 | 0.09 | 176 | 229 | 124 | 15 |
| 2019 | 42 | 0.20 | 0.21 | 386 | 514 | 259 | 17 |
| 2020 | 42 | 0.16 | 0.16 | 320 | 412 | 227 | 15 |
| 2021 | 42 | 0.14 | 0.13 | 270 | 347 | 193 | 15 |
| 2022 | 42 | 0.13 | 0.12 | 259 | 332 | 185 | 14 |

*No TV survey from 2007 to 2010.

Table 19.3.1. Nephrops in FU19 (SW and SE Ireland). Forecast inputs (bold) and historical estimates of mean weight in landings and harvest rate (landings + dead discards)/(abundance estimate), discard rate (discards divided by landings + discards) and dead discard rate as dead discards divided by removals (landings + dead discards). * 2019 revision due to code error.

| Year | Landings in number | Total discards* in number | Removals in number | Discard Rate number | Dead discard rate number | UWTV abundance estimate | 95\% Conf. intervals | Harvest rate | Landings | Total discards* | Mean weight in landings | Mean weight in discards |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions | millions | millions | \% | \% | millions | millions | \% | tonnes | tonnes | grammes | grammes |
| 2006 | 26.2 | 2.6 | 28.1 | 8.9 | 6.8 | na | na | na | 741 | 37 | 28.3 | 14.4 |
| 2007 | 30.8 | 1.5 | 31.9 | 4.8 | 3.6 | na | na | na | 957 | 26 | 31.1 | 17 |
| 2008 | 25.2 | 5.4 | 29.3 | 17.7 | 13.9 | na | na | na | 851 | 105 | 33.7 | 19.4 |
| 2009 | 28.4 | 18.5 | 42.3 | 39.5 | 32.8 | na | na | na | 868 | 269 | 30.5 | 14.5 |
| 2010 | 23.2 | 19.0 | 37.4 | 45.1 | 38.1 | na | na | na | 687 | 257 | 29.6 | 13.5 |
| 2011 | 25.8 | 32.4 | 50.1 | 55.7 | 48.5 | 665 | 171 | 7.5 | 643 | 409 | 24.9 | 12.6 |
| 2012 | 32.3 | 37.3 | 60.2 | 53.6 | 46.4 | 594 | 111 | 10.1 | 849 | 473 | 26.3 | 12.7 |
| 2013 | 29.5 | 36.5 | 56.8 | 55.3 | 48.1 | 487 | 161 | 11.7 | 794 | 436 | 26.9 | 11.9 |
| 2014 | 16.3 | 11.4 | 24.9 | 41.1 | 34.4 | 636 | 188 | 3.9 | 468 | 161 | 28.6 | 14.1 |
| 2015 | 17.0 | 11.8 | 25.9 | 41.1 | 34.3 | 482 | 126 | 5.5 | 507 | 167 | 29.8 | 13.8 |
| 2016 | 19.7 | 13.6 | 29.9 | 40.8 | 34.1 | 399 | 99 | 7.5 | 590 | 193 | 29.9 | 14.2 |
| 2017 | 14.6 | 9.6 | 21.8 | 39.7 | 33.1 | 499 | 120 | 4.4 | 420 | 139 | 28.8 | 14.5 |
| 2018 | 8.4 | 4.5 | 11.8 | 34.8 | 28.6 | 176 | 53 | 6.7 | 238 | 71 | 28.2 | 15.7 |
| 2019* | 9.1 | 8.5 | 15.4 | 48.2 | 41.1 | 386 | 127 | 4.0 | 249 | 112 | 27.4 | 13.3 |
| 2020 | 9.7 | 10.1 | 17.2 | 51 | 43.9 | 320 | 93 | 5.4 | 249 | 136 | 25.8 | 13.5 |
| 2021 | 14.6 | 12.7 | 24.1 | 46.5 | 39.4 | 270 | 77 | 8.9 | 415 | 173 | 28.5 | 13.6 |
| 2022 |  |  |  |  |  | 259 | 73 |  |  |  |  |  |
| Average 2019-2021 |  |  |  | 48.6 | 41.5 |  |  |  |  |  | 27.2 | 13.5 |



Figure 19.1.1. Nephrops in FU19 (Ireland SW and SE Coast). Revised discrete patches overlaid on overlaid on proportion of Nephrops in the Irish landings overlaid on international OTB effort (red=0\% Nephrops; blue=50-60\% Nephrops; grey=unknown (no Irish landings).


Figure 19.1.2. Nephrops in FU19 (Ireland SW and SE Coast). Time-series of the number of Irish vessels reporting landings of Nephrops from FU19 with a >10 threshold.


Figure 19.1.3. Nephrops in FU19 (Ireland SW and SE Coast). Combined box and kite plot of vessel power by year. The blue line indicates the mean.


Figure 19.2.1. Nephrops in FU19 (Ireland SW and SE Coast). Landings in tonnes by country.


Figure 19.2.2. Nephrops in FU19 (Ireland SW and SE Coast). Trawl effort for Irish OTB vessels where >30\% of landed weight was Nephrops.


## Length frequencies for catch (dotted) and landed(solid): Nephrops in FU19



Figure 19.2.4. Nephrops in FU19 (Ireland SW and SE Coast). Mean size trends for catches (dotted) and whole landings (solid) by sex 2002-2021. Vertical lines displayed are Minimum Conservation Reference Size $\mathbf{2 5} \mathbf{~ m m}$ Carapace Length (CL) and 35 mm CL.


Figure 19.2.5. Nephrops in FU19 (Ireland SW and SE Coast). Annual sex ratio of landings (2008-2020) and catch (20082021).


Figure 19.2.6. Nephrops in FU19 (Ireland SW and SE Coast). Mean weight in catch data for all grounds in FU19 by sex with loess smoother and showing cyclical trends.


Figure 19.2.7. Nephrops in FU19 (Ireland SW and SE Coast). Annual estimated mean weights (gr) in the landings and discards.


Figure 19.2.8. Nephrops in FU19 (Ireland SW and SE Coast). Violin and box plot a of adjusted burrow density (burrow/m²) distributions by year from 2006-2022. The blue line indicates the mean density over time. The horizontal black line represents the median, white box is the interquartile range, the black vertical line is the range and the black dots are outliers. No estimate available for Galley Ground 4 in 2011, Galley Ground 1 in 2012. No TV survey from 2007 to 2010.


Figure 19.2.9. Nephrops in FU19 (Ireland SW and SE Coast). Time-series of total abundance estimates for FU19 (error bars indicate $95 \%$ confidence intervals) and $\mathrm{B}_{\text {trigger }}$ is dashed line.

Length frequencies for catch (dotted) and landed(solid): Nephrops in FU 19 IGFS-WIBTS-Q4 [G7212]


Figure 19.2.10. Nephrops in FU19 (Ireland SW and SE Coast). Mean size trends for catches by sex from IGFS- WIBTS-Q4 [G7212] Irish survey 2003-2021. Vertical lines displayed are Minimum Conservation Reference Size 25 mm Carapace Length (CL) and 35 mm CL.


Figure 19.3.1. Nephrops in FU19 (Ireland SW and SE Coast). Harvest Rate (\% dead removed/UWTV abundance). The dashed and solid lines are the MSY proxy and the harvest rate respectively.

# 20 Norway lobster (Nephrops novegicus) in divisions 7.g and 7.h, Functional Units 20 and 21 (Celtic Sea) 

## Type of assessment in 2022

A full UWTV based assessment was carried out and catch options based on the stock-specific FMSY reference point estimated by WGCSE 2016 using the methods applied to other Nephrops stocks at WKFMSYREF4 (ICES, 2016) and a newly proposed MSY Btrigger estimate (ICES, 2021a; Annex 3). This stock assessment is available in the ICES Transparent Assessment Framework (TAF) here.

## ICES advice applicable to 2021

"ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, catches in 2021 that correspond to the $F$ ranges in the MAP are between 1682 tonnes and 1710 tonnes, assuming recent discard rates. The entire range is considered precautionary when applying the ICES advice rule.

To ensure that the stock in functional units 20 and 21 is exploited sustainably, management should be implemented at the level of the combined functional units 20 and 21."

## ICES advice applicable to 2022

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2018-2020, catches in 2022 should be no more than 1978 tonnes.

To ensure that the stock in functional units (FUS) 20-21 is exploited sustainably, management should be implemented at the level of the combined FU 20-21.

ICES notes the existence of a management plan, developed and adopted by one of the relevant management authorities for Subarea 7. ICES considers this plan to be precautionary when implemented at the FU level."

### 20.1 General

## Stock description and management units

The FU20-21 Nephrops stock is included in the whole ICES Area 7 together with Irish Sea East and West [FU14, FU15], Porcupine Bank [FU16], Aran Islands [FU17], northwest Irish Coast [FU18], southeast and southwest Irish Coast [FU19], Smalls [FU22]. The TAC is set for Subarea 7 which does not correspond to the stock area.

Historically FU20-22 fishery and sampling data covered an amalgamation of several spatially distinct mud patches; FU20 NW Labadie, Baltimore and Galley, FU21 Jones and Cockburn and FU22 the Smalls. WGCSE 2013 recommended that FU20-22 should be split into FU20-21 combined and FU22 for the purposes of assessment and advice provision. There is evidence that the Celtic Sea Nephrops patches are linked in meta-population sense (O'Sullivan et al., 2015). However, fishing mortality and biological parameters (density, growth, $M$, etc.) may vary across the different patches. The map below shows FU20-21 assessment area (blue) and TAC area (red). There is no evidence that the individual functional units belong to the same stock. See Section 18 for details on Nephrops in Subarea 7 general section.


## Ecosystem aspects

Details of the ecosystem on FU20-21 are provided in the stock annex updated by WKCELT.

## Fishery description

Ireland, France and the UK are the main countries involved in the FU20-21 Nephrops fishery. In the early 2000s the Republic of Ireland fleet had on average less than $10 \%$ of the landings and this has increased to over $80 \%$ from this FU in recent 3 year period. A description of this fleet is given in the stock annex. The fishery on FU20-21 grounds operates throughout the year with a seasonal trend, weather permitting, and has expanded in the mid-2000s. The time-series of numbers of vessels with landings greater than 10 tonnes is updated in Figure 20.1.1. The time-series of vessel power is shown as a box and kite plot in Figure 20.1.2. In recent years the Irish fleet have increased landings from the southern part of the grounds (see stock annex). Recently several newer vessels specializing in Nephrops fishing have participated periodically in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates. There has been a trend for Irish vessels to switch to multi (quad) rig trawls since 2012. These vessels are more efficient at catching Nephrops (BIM, 2015).

French trawlers targeting Nephrops in the Celtic Sea operate mainly in the FU20-21. France dominated in the landings in the early 2000s on average $90 \%$ of landings and this has decreased to about $14 \%$ in recent times (2019-2021). A description of this fleet is given in the stock annex.
There is an increase in participation by the UK in this fishery in the most recent years. The UK fleet had on average $4 \%$ of the landings from this FU in recent times (2019-2021) with highest landings recorded in 2018 (411 t).

## Fishery in 2021

## Ireland

In 2021, 53 Irish vessels reported landings from FU2021. Of these, 32 vessels reported landings in excess of 10 t which is a significant reduction compared to previous years

Landings by metier is quite mixed compared to other Nephrops fishery, however, since 2020 there was a significant decrease in landings by metier OTB_CRU-100_119 as shown in Figure 20.1.3 with the majority of landings now taken by the OTB_DEF_100-119 metier. The reason for this is not known and will be investigated.

## France

In 2021, 29 French vessels reported landings from FU20-21 where many of these switch between FU20-21 and FU22 within a trip. Of these 5 vessels reported landings in excess of 10 t .

## UK

10 UK (E\&W) vessels reported landings for FU20-21.

## Information from stakeholders

None presented.

### 20.2 Data

## InterCatch

Data were available in InterCatch and used for catch data only. French data were provided directly by the national expert and not extracted from InterCatch.

## Landings

The reported landings time-series is shown in Figure 20.2.1 and Table 20.2.1.
The reported Irish landings from FU20-21 have increased since the mid-2000s to the highest in the Irish time-series in 2019 ( 2219 t). In 2020 and 2021 Irish landings were the lowest values reported since 2005. French landings have gradually decreased since the early 2000s to the present reported landings of 114 t . Reported landings from the UK have fluctuated with an increasing trend since 2015. There was revision to the 2019 UK(E\&W) landings from 551 t to 276 t due to a code error and these revised data are now used. Minor landings were reported by the UK ( 9 t ) and Belgium ( 3 t ) in 2021.

The overall fishing profile remains typically seasonal with the majority of the Irish and UK landings coming from the second quarter (see stock annex).

## Effort

Effort data are available for the Irish Nephrops directed fleet in FU2021 from 1995-2021. The effort series is based on the same criteria for FU15, 16, 17, 19and 22 ( $30 \%$ landings threshold) and will be contingent on the accuracy of landings data reported in logbooks. Effort data are not standardized, and hence do not take into account vessel capabilities, efficiency, seasonality or other
factors that may bias perception of lpue as an abundance trend over the longer term. These data are not used in the assessment.

WGCSE 2015 recommended that effort data in Kw days should be presented as these data are more informative than effort data uncorrected for vessel power. Effort data are available from 1995 for the Irish otter trawl Nephrops directed fleet. In 2021, this fleet accounted for $\sim 90 \%$ of the Irish landings compared with an average of $70 \%$ over the time period. Effort shows a generally increasing trend since the mid-2000s with a sharp decrease between 2015 and 2017 and also since 2020. (Figure 20.2.2 and Table 20.2.2).

Effort data in KW days are not available for France. Previously effort data were reported from 1983 to 2008 for the French Nephrops fleet for the combined Celtic Sea FU20-22 (see stock annex). Since 2009, a new registration system of official French statistics has changed the way fishing effort is computed and a new threshold method of 500 kg landed by trip is used to report effort. French fishing effort reported in hours and lpue ( $\mathrm{kg} / \mathrm{hr}$ ) since 2009 shows an overall declining trend (Table 20.2.3). Data not reported for 2021.

## Sampling levels

Sampling levels, data aggregating and raising procedures were reviewed by WKCELT 2014 and are documented in the stock annex. The time-series of sampling levels is shown in Table 20.2.4, and remains sparse due to the offshore nature of the fishery although good progress is being made by Ireland in recent years.
There was a revision to the Ireland 2019 sampling data set due to the inclusion of a valid sample that was discovered as a result of QA process of a SQL server migration. The inclusion of this sample data to the 2019 Ireland fishery data summary had a minor impact on the assessment summary. The details data of this revision was presented to the WGCSE 2020 meeting and was accepted.

## Commercial length-frequency distributions

Prior to 2012 there was insufficient Irish sampling to generate length-frequency distributions although since then sampling levels have improved. For France limited data were available for 1997 and 2010-2013. In 2019 sampling data was not used due to quality issues (see stock annex for details). In 2020 one sample was available but not deemed useful for assessment purposes. In 2021 two samples were available but again not deemed useful for assessment.
Length-frequency distributions of landings and discards for both countries from 2012 to 2021 are presented in Figure 20.2.3 along with the European minimum conservation reference size ( 25 CL mm ) and French ( 35 CL mm ) minimum landings size also shown. In 2019 France provided sample data numbers and raised data, however, it was not included in the assessment this year due to data quality issues. In 2020 sampling data were not available from France due to the COVID-19 pandemic and also the quite low level of participation in the fishery.
In 2020 there is a lack of small individuals in the catch in 2020 from Irish sampling the reason for this is unclear at present.

The short series on LFDs for both countries shows that the LFDs differ between the two countries. A higher proportion of the French catch consists of large individuals ( $>35 \mathrm{~mm}$ ) - on average $70 \%$ compared to $41 \%$ for the Irish fishery for the available comparable time-series.

## Sex ratio

The sex ratio is male biased from the available French and Irish sampling data (Table 20.2.5).

## Mean weight explorations

The French dataset provided to WGCSE 2017 (years 2012-2015) results in an increase in mean weights and decrease in removals from that previously reported at WGCSE 2016 (Table 20.2.6). The working group accepted the French dataset, and this is used to calculate the estimated annual mean weights in the landings and discards.
The length-weight relationship as described in stock annex is used to raise both countries sampling data, which are based on Scottish data (Pope and Thomas, 1955).
The mean weight in the landings for France is higher than that in the Irish landings (Table 20.2.7). The estimated annual mean weights in the landings and discards by country and also combined scaled to the international landings is shown in Table 20.2.8 and Figure 20.2.4). There is a big decline in 2016 to 2017 which coincides with the very high UWTV estimate of abundance - which could indicate a strong year class.

## Discards

For the Irish data, discard rates have been estimated using unsorted catch and discards sampling. This involves unsorted catch and discard samples being provided by vessels or collected by observers at-sea on discard trips. The catch sample is partitioned into landings and discards using an on-board discard selection ogive derived for the discard samples. Due to sparse sampling effort, annual data are used to derive length distributions and selection ogives. Figure 20.2.5 shows the annual discard ogive from the Irish sampling used to partition the catch. The lack of smaller individuals was also evident in the 2020 discard ogive. The length-weight regression parameters given in the stock annex are used to calculate sampled weights and appropriate annual raising factors. The sampling intensity and coverage has varied over the short time-series, and is relatively poor compared to other Nephrops stocks, but is considered adequate for stock assessment purposes.

Estimated discard rates range between 12-41\% of total catch by number and $7-27 \%$ of total catch by weight in the Irish fishery shown in Table 20.2.7. The 2020 discard rates could be related to a change in this fishery mainly comprised of OTB_DEF metiér, however, this will be investigated further. In the French fishery estimated discard rates range between $25-78 \%$ of total catch by number and 16-56\% of total catch by weight shown in Table 20.2.6.

Estimated discard rates for both countries combined in shown in Table 20.2.8 and these range between $24-52 \%$ of total catch by number and $14-31 \%$ of total catch by weight. Discard rate of females tends to be higher due to the smaller average size and market reasons as is observed in other Nephrops fisheries.

There is no information on discard survival rate in this fishery. $25 \%$ is assumed in line with other Nephrops stocks in the Celtic Sea (Charuau et al., 1982).

Gear selectivity trials by Bord Iascaigh Mhara (BIM, 2017) reported a $64 \%$ survivor rate for Nephrops caught in a trawl with a SELTRA selectivity device in the outer Galway Bay area.

Table 20.3.1 gives weights, numbers and mean weights of the landings and discard raised internationally according to the stock annex.

## Abundance indices from UWTV surveys

The methods used during the survey were similar to those usually employed for UWTV surveys [U5917] of Nephrops stocks around Ireland and elsewhere and are documented by WKNEPHTV (ICES, 2007), WKNEPHBID (ICES, 2008), SGNEPS (ICES, 2009; 2010; 2012), WGNEPS (ICES, 2013; 2014; 2015; 2016a; 2017; 2018a, 2020, 2021, 2022), WKNEPS (ICES, 2016b; 2018b), Leocádio, A., et al, 2018 and Dobby H., et al, 2021.

SGNEPS (ICES, 2012) recommended that a CV (or relative standard error) of $<20 \%$ is an acceptable precision level for UWTV survey estimates of abundance. UWTV surveys conducted in 2012 are deemed exploratory as stations were chosen based on areas heavily fished by vessels (Doyle et al., 2013). These are likely to give biased estimates of density and cannot be extrapolated to estimate density for the whole area. A randomised isometric grid design was employed with UWTV stations at 6.0 nmi intervals for 2013-2021 surveys.

A review of the kriging analyses by two different software packages for survey years 2013 and 2014 was investigated as part of the transition of the assessment to the ICES TAF process This was reviewed by an external expert. The results from SURFER and RGeostats software were very close and full details are available on the ICES TAF stock GitHub repo and also in Annex 3 (ICES, 2021a). The summary statistics from the RGeostats software are now used in the assessment for those years. The 2013 survey achieved partial coverage $\sim 60 \%$ of the total area. The 2013 abundance has been scaled up to the entire area since densities in the un-surveyed part of the ground were not significantly different in 2014. From 2014 to 2022 full survey coverage was achieved. The geo-statistical analysis for years 2013 to 2022 follows the steps documented in Doyle et al., 2022.

The 2022 mean burrow density was 0.10 burrows $/ \mathrm{m}^{2}$ compared with 0.12 burrows $/ \mathrm{m}^{2}$ in 2021. The 2022 geostatistical abundance estimate was 1032 million a $14 \%$ decrease on the abundance for 2021 with a CV of $5 \%$ which is well below the upper limit of $20 \%$ recommended by SGNEPS 2012. There was a slight decrease in densities observed in 2022. Figure 20.2 .6 shows the krigged contour and density plots for the time-series. The summary statistics from this geostatistical analysis are given in Table 20.2.9 and plotted in Figure 20.2.7. The geostatistical abundance estimate adjusted is derived using the mean of the krigged grid, where the mean of the observations is reported in Table 20.2.9. The estimation variance of the survey is very low (CVs in the order 5\%).

## Groundfish survey data

There are two IBTS-GFS catching Nephrops in FU20-21: French groundfish survey EVHOE-WI-BTS-Q4 [G9527] since 1997 and Irish groundfish survey-Q4: IGFS-WIBTS-Q4 [G7212] commenced in 2003 (Stokes et al., 2014). These provide information on length-frequency compositions, mean size in the catches, cpue of Nephrops in FU20-21 (ICES, 2015). The mean size of the catches is stable over the time-series except in 2006 and 2008 which signals recruitment into the fishery in 2006 and 2007 as shown by the Irish IBTS survey in Figure 23.2.8 and the French IBTS survey (Figure 23.2.9). There is also a signal of recruitment in 2018 mean size from IGFS survey. There is no 2017 length dataset for EVHOE due to research vessel breakdown.

### 20.3 Assessment

## Comparison with previous assessments

The WGCSE 2021 carried out a full UWTV based assessment for this stock using the stock-specific reference points were estimated by the 2016 working group based on methods for other

Nephrops stocks used by WKMSYREF4 (ICES, 2016). This is in accordance with recommendations by WKCELT 2014 where data improvements have been made for this stock such as:

- complete survey coverage of the stock area giving quality assured density estimates and abundance estimates conforming to WGNEPS recommendations; and also
- improved sampling data achieving better coverage and robust estimates of the various parameters need to calculate catch options (e.g. mean weight in the landings and discards, discard percentage in numbers).
- Proposal of MSY Btrigger based on seven years of survey data.


## State of the stock

UWTV abundance estimates suggest that the stock size has fluctuated over the time-series. The 2022 estimate is a decrease from 2021 estimate by $14 \%$.

The 2022 estimate is above the newly proposed MSY $B_{\text {trigger }}(450$ million). The 2022 estimate (1032 million) is below the average of the series (geomean [2014-2022]: 1610 million).

Table 23.3.1 and Figure 23.3.1 summarize recent harvest rates which have been below the Fmsy proxy except in 2019 where the harvest rate is $19.2 \%$ which is a result of the low stock abundance estimate and high catches.

### 20.4 Catch scenario table

Catch scenario table inputs and estimates of mean weight in landings and harvest ratios are presented in Table 23.3.1 and summarised below.

In line with previous practice an average (2019-2021) of mean weights is used to account for this variability. Three year average (2019-2021) of proportion of removals retained was used as is standard for other Nephrops stocks.

The basis for the catch scenario:

| Variable | Value | Notes |
| :--- | :---: | :--- |
| Stock abundance (2023) | 1032 | Numbers of individuals (millions); UWTV survey 2022 |
| Mean weight in projected landings | 30.3 | Average 2019-2021 in grammes |
| Mean weight in projected discards | 16.2 | Average 2019-2021 in grammes |
| Projected discards | 17.3 | Proportion by number; Average 2019-2021 |
| Discards survival | 25 | Proportion by number |
| Projected dead discards | 13.7 | Proportion by number; Average 2019-2021 |

A prediction of landings for the FU20-21 using the approach agreed procedure proposed at WKNEPH 2009 and outlined in the stock annex will be made on the basis of the 2022 UWTV survey. This will be presented in October 2022 for the provision of advice.

### 20.5 Reference points

New reference points were estimated by WGCSE 2016 using the same method and approach used at WKMSYREF4 (ICES, 2016). The detailed analysis is available in working document 11
(WGCSE, 2016). In the case of FU20-21 there is a limited number of years for which lengthfrequency data were available, so the three-year moving window could only be applied to give two estimates. The resulting potential $\mathrm{F}_{\mathrm{mSY}}$ harvest rates and ranges are given in the following table.

| YEAR | FMAX | FMAX.LOW | FMAX.UP | F35 | F35.LOW | F35.UP | F0.1 | F0.1.LOW | F0.1.UP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 9.12 | 6.51 | 12.60 | 11.03 | 6.11 | 13.21 | 5.91 | 5.08 | 15.11 |
| 2013 | 9.45 | 6.71 | 13.26 | 11.17 | 6.30 | 13.78 | 6.10 | 5.23 | 15.93 |

Given the low density in the area and combined sex $\mathrm{F}_{0.1}$ was considered and appropriate $\mathrm{F}_{\mathrm{MSY}}$ proxy.

| STOCK CODE | MSY FLOWER* | FMSY* | MSY FUPPER*WITH AR | MSY BTRIGGER | MSY FUPPER*WITH NO AR |
| :--- | :---: | :---: | :---: | :---: | :---: |
| nep-2021 | $5.9 \%$ | $6.0 \%$ | $6.0 \%$ | Not defined | $6.0 \%$ |

* Harvest rate (HR).

At WGCSE 2021 MSY Btrigger estimate was proposed using the same method and process used at WKMSYREF4 (ICES, 2016). The detailed analysis is available and was externally reviewed (ICES, 2021a; Annex 3). The estimate was based data on survey years 2014 to 2021 excluding year 2017 and value is given in table below:


### 20.6 Management plans

There is no specific management plan for the FU 20-21 Nephrops.
The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to Norway lobster (Nephrops norvegicus) by functional unit in ICES subarea 7 and also demersal stocks. There is currently no agreement with the UK regarding this plan.

### 20.7 Quality of assessment and forecast

Since the benchmark in 2014 UWTV and sampling coverage has been improving in this area. There are now nine years of full UWTV survey coverage (2014-2022). Since 2019 the survey camera system and reviewing method changed where a new HD system is used (ICES, 2019). A comparison showed no significant difference in density estimates between the new (HD) and the old method (SD) for FU 2021 (ICES, 2022b). Previous assumptions relating to correction factors are still applied for this FU 20-21.

There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008; SGNEPS 2009; WGNEPS 2014 and Dobby et al,
2021). Ultimately there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate, although no more precise (WKNEPH, 2009). The survey estimates themselves are very precisely estimated (CVs $\sim 5 \%$ ) given the homogeneous distribution of burrow density and the modelling of spatial structuring. The cumulative bias estimates for FU20-21 are largely based on expert opinion. The precision of these bias corrections cannot yet be characterised, but is likely to be lower than that observed in the survey.

At WGCSE 2018, the group recommended that a review of historical survey data should be undertaken given the large fluctuations observed in the short time-series to date for this survey, that is, to randomly check $20 \%$ of UWTV stations in years 2016 and 2017. This process was conducted in July 2018 during the FU20-21 UWTV survey. The analysis was presented to WGNEPS (2018a) and subsequently to the 2019 WGCSE meeting where full details are available in R-markdown (ICES, 2018a; Annex 7). Results are briefly summarised here. The analyses showed a low increase in the review counts for 2016 stations comparing them with the survey counts $(3.8 \%$ increase), and a high decrease in the review counts for 2017 stations comparing them with the survey counts ( $30.8 \%$ decrease). Next the review count data were swapped with the survey count data and abundance was calculated for both years using the "RGeostats" package (Renard D. et al., 2015), following the same procedure that was carried out in those years previously. The geostatistical results showed an increase of $4.6 \%$ in 2016 abundance estimate (from 1879 million to 1966 million), and a decrease of $4 \%$ in 2017 abundance estimate (from 4428 million to 4250 million). The geo-statistical CVs were in the order of $3.7 \%$ to $4.4 \%$, which are well below the upper limit recommendation of $20 \%$ (ICES, 2012).

Following this analysis WGNEPS 2018 recommended to include guidelines on quality control where there are large unexplained fluctuations between abundance estimates from previous years in the manual for Nephrops underwater TV surveys (Dobby et al., 2021). In that it is recommended to review $20 \%$ of the survey stations, and when the partial review differs more than $20 \%$ from the survey counts, then a full review of the survey should be considered.

These were also followed in 2019 given the substantial decrease observed. A random selection of $20 \%$ of UWTV stations were reviewed. Full details are available in R-markdown (ICES, 2019). The results showed an overall increase in the review counts for these selected stations comparing them with the survey counts ( $15.5 \%$ increase). This process confirmed the observed low density estimates which are used to calculate the abundance estimate for determining catch scenarios for 2020.

Sampling of landing and discards for FU20-21 remains low but there is a limited number of years for which length-frequency data were available so the three year moving window could only be applied to give two estimates to calculate $\mathrm{F}_{\text {MSY }}$ reference points.

French and Irish trawlers cover different areas and have presented contrasting features over the last decade. The French fleet moved gradually from the "Smalls" Ground (mainly 31E3) to the "Labadie" (30E2, increase of 28E2 in the early 2010s, although no trend is revealed within FU2021 throughout the overall time-series): in the late 1990s, more than $40 \%$ of French landings were reported from the "Smalls" area whereas by the end of 2000s the contribution of this rectangle became minor (less than 10\%). Irish vessels have increased their production on FU20-21 since the mid-2000s and a gradual expansion towards the southern rectangles is obvious during the recent years (stock annex).

### 20.8 Recommendations for next benchmark

This stock was last benchmarked by WKCELT (ICES, 2014). WGCSE will keep the stock under close review and recommend future benchmark as required.

### 20.9 Management considerations

The indications are the Nephrops in FU20-21 are well exploited now relative to the past. Overall effort in the French fishery has declined to less than $25 \%$ of the peak effort observed in the early 1990s whereas there has been a big increase in Irish effort over the recent years with sharp decline in 2020.

Overall the Irish fishery in the area expanded with the exception of 2020, whereas the French fishery continued to decline. The fishing patterns of the French and Irish fleet are very different with the Irish fleet specialising on Nephrops whereas the French fishery remains more mixed. French Nephrops fisheries in this area are fairly mixed also catching whiting, cod, megrim, anglerfish and other demersal species (Davie and Lordan, 2011). Nephrops tend to dominate the landings of Irish fisheries in the area but catches are more mixed in the North ( $\sim 50 \%$ Nephrops) and cleaner Nephrops towards the south ( $\sim 75 \%$ Nephrops) (Gerritsen et al., 2012). The French trawlers showed an overall decline in effort and landings during the last decade, mainly explained by decommissioning schemes associated with constraints linked to fuel prices.
In recent years several newer vessels specializing in Nephrops fishing have participated in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates.

From 2016, fisheries catching Nephrops in Subarea 7 are covered by the EU landings obligation (EU, 2015). A high survivability exemption applies to creel fisheries from the landings obligation. Irish discard survival experiments indicate that the trawl discard survival may be around $64 \%$ (BIM, 2017). As a result, an exemption from the landings obligation based on high survivability has been granted by the European Commission. The average discard rate by weight for FU2021 over the last three years is $10 \%$. Catch advice and scenarios are provided this year on the assumption that discarding is assumed to continue at the recent average.

UWTV survey coverage has improved. A new survey point available by autumn 2022 provides a more up to date estimate of density and abundance. The most up to date survey information is used as an abundance estimate for this stock.

Landings data are adjusted to take into account landings that have been misreported from FU16 since 2011. This adjustment is thought to be reasonably accurate (See Section 19).

ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES division level. Management at the functional unit level could provide controls to ensure effort and catch were in line with resources available.

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Table 20.2.1. Nephrops FU 20-21. Landings in tonnes by country.

| Year | FU 20-21 Landings ( $\mathbf{t}$ ) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | France | Rep. of Ireland | UK | Belgium | Total |
| 1995 | 3419 | 117 | na |  | 3536 |
| 1996 | 2721 | 101 | na |  | 2822 |
| 1997 | 1957 | 81 | na |  | 2038 |
| 1998 | 1583 | 130 | na |  | 1713 |
| 1999 | 1051 | 83 | 18 |  | 1152 |
| 2000 | 1661 | 107 | 10 |  | 1778 |
| 2001 | 1750 | 69 | 14 |  | 1833 |
| 2002 | 2559 | 104 | 11 |  | 2674 |
| 2003 | 2796 | 148 | 9 |  | 2953 |
| 2004 | 2140 | 299 | 4 |  | 2443 |
| 2005 | 2008 | 455 | 6 |  | 2469 |
| 2006 | 2066 | 450 | 7 |  | 2523 |
| 2007 | 1816 | 600 | 3 |  | 2419 |
| 2008 | 2036 | 937 | 7 |  | 2980 |
| 2009 | 1930 | 1202 | 13 |  | 3145 |
| 2010 | 975 | 756 | 62 |  | 1793 |
| 2011 | 566 | 637 | 34 |  | 1237 |
| 2012 | 453 | 708 | 28 |  | 1189 |
| 2013 | 486 | 844 | 57 |  | 1387 |
| 2014 | 465 | 1342 | 29 |  | 1837 |
| 2015 | 355 | 1620 | 141 |  | 2116 |
| 2016 | 477 | 1531 | 445 |  | 2453 |
| 2017 | 341 | 1113 | 395 | 0.2 | 1849 |
| 2018 | 195 | 1197 | 411 | 0.2 | 1803 |
| 2019 | 218 | 2219 | 286 | 0.1 | 2723 |
| 2020 | 75 | 336 | 2 | 0.03 | 413 |
| 2021 | 114 | 610 | 9 | 3 | 736 |

Table 20.2.2. Nephrops FU 20-21. Effort data for the Irish otter trawl Nephrops directed fleet. Effort for vessels where 30\% of the landed weight was Nephrops.

| Year | Effort ('000's KwDays) | Landings (tonnes) |
| :---: | :---: | :---: |
| 1995 | 57 | 104 |
| 1996 | 49 | 74 |
| 1997 | 40 | 59 |
| 1998 | 56 | 102 |
| 1999 | 37 | 48 |
| 2000 | 39 | 62 |
| 2001 | 29 | 45 |
| 2002 | 78 | 165 |
| 2003 | 82 | 86 |
| 2004 | 159 | 164 |
| 2005 | 255 | 360 |
| 2006 | 301 | 348 |
| 2007 | 402 | 512 |
| 2008 | 562 | 920 |
| 2009 | 801 | 1,249 |
| 2010 | 498 | 633 |
| 2011 | 424 | 535 |
| 2012 | 357 | 534 |
| 2013 | 445 | 672 |
| 2014 | 885 | 1,170 |
| 2015 | 1,180 | 1,542 |
| 2016 | 920 | 1,404 |
| 2017 | 704 | 1,004 |
| 2018 | 695 | 1,084 |
| 2019 | 1,185 | 2,153 |
| 2020 | 184 | 245 |
| 2021 | 342 | 443 |

Table 20.2.3. Nephrops FU 20-21. Effort data for the French fleet. * data not available.

| Year | Effort France ('000 hrs) | Lpue France (kg/hr) |
| :---: | :---: | :---: |
| 1983 | 231 | 14 |
| 1984 | 205 | 16 |
| 1985 | 203 | 16 |
| 1986 | 163 | 15 |
| 1987 | 190 | 15 |
| 1988 | 171 | 16 |
| 1989 | 179 | 17 |
| 1990 | 230 | 16 |
| 1991 | 225 | 11 |
| 1992 | 277 | 12 |
| 1993 | 268 | 13 |
| 1994 | 259 | 14 |
| 1995 | 239 | 15 |
| 1996 | 220 | 14 |
| 1997 | 187 | 13 |
| 1998 | 155 | 13 |
| 1999 | 151 | 11 |
| 2000 | 194 | 14 |
| 2001 | 170 | 15 |
| 2002 | 166 | 19 |
| 2003 | 192 | 18 |
| 2004 | 153 | 16 |
| 2005 | 147 | 16 |
| 2006 | 137 | 16 |
| 2007 | 102 | 19 |
| 2008 | 100 | 23 |
| 2009 | 93 | 23 |
| 2010 | 67 | 17 |


| Year | Effort France ('000 hrs) | Lpue France (kg/hr) |
| :---: | :---: | :---: |
| 2011 | 52 | 12 |
| 2012 | 42 | 13 |
| 2013 | 48 | 12 |
| 2014 | 36 | 15 |
| 2015 | 35 | 11 |
| 2016 | 35 | 15 |
| 2017 | 34 | 11 |
| 2018 | 21 | 10 |
| 2019 | 22 | 11 |
| 2020 | 12.5 | 6.1 |
| 2021* |  |  |

Table 20.2.4.a. Nephrops FU 20-21. Sampling levels by Ireland.

| IRELAND |  | Number of Samples |  | Numbers Measured |  |  | Landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Quarter | Catch | Discards | Landings | Catch | Discards |  |
| 2009 | 2 | 1 | 0 |  | 489 | 0 |  |
| 2010 | 2 | 1 | 0 |  | 461 | 0 |  |
| 2011 | 2 | 1 | 0 |  | 270 | 0 |  |
| 2012 | 1 | 8 | 5 | 1 | 2,654 | 2,024 | 1,747 |
| 2013 | 1 | 1 | 1 |  | 319 | 423 |  |
| 2013 | 2 | 9 | 7 | 1 | 2,514 | 2,038 | 2,187 |
| 2014 | 2 | 2 | 2 |  | 718 | 782 |  |
| 2015 | 1 | 0 | 0 | 1 | 0 | 0 | 1,724 |
| 2015 | 2 | 6 | 6 | 2 | 2,714 | 3,997 | 3,204 |
| 2015 | 3 | 0 | 0 | 4 | 0 | 0 | 4,750 |
| 2015 | 4 | 2 | 2 |  | 650 | 419 |  |
| 2016 | 2 | 8 | 5 | 1 | 2,859 | 1,485 | 384 |
| 2016 | 4 | 3 | 2 | 4 | 767 | 1,678 | 1,743 |
| 2017 | 1 | 2 | 1 | 1 | 722 | 297 | 1,616 |
| 2017 | 2 | 7 | 4 | 1 | 2,813 | 1,035 | 365 |
| 2017 | 3 | 3 | 1 |  | 1,154 | 296 |  |
| 2017 | 4 | 12 | 7 |  | 3,631 | 1,983 |  |
| 2018 | 1 | 3 | 3 |  | 987 | 1,036 |  |
| 2018 | 2 | 17 | 17 |  | 6,691 | 5,742 |  |
| 2018 | 3 | 2 | 0 |  | 389 | 0 |  |
| 2018 | 4 | 2 | 1 |  | 544 | 369 |  |
| 2019 | 1 | 8 | 6 |  | 2,691 | 3,103 |  |
| 2019 | 2 | 12 | 10 |  | 4318 | 3,738 |  |
| 2019 | 3 | 1 | 1 |  | 373 | 520 |  |
| 2019 | 4 | 0 | 0 |  | 0 | 0 |  |

Table 20.2.4.a. Nephrops FU 20-21. Sampling levels by Ireland.

| IRELAND |  | Number of Samples |  |  | Numbers Measured |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Quarter | Catch | Discards | Landings | Catch | Discards | Landings |
| 2020 | 1 | 11 | 9 |  | 3,412 | 1,934 |  |
| 2020 | 2 | 10 | 8 |  | 3,581 | 2,448 |  |
| 2020 | 3 | 2 | 0 |  | 689 | 0 |  |
| 2020 | 4 | 0 | 0 |  | 0 | 0 |  |
| 2021 | 1 | 9 | 5 |  | 2,987 | 1,495 |  |
| 2021 | 2 | 14 | 10 |  | 5,991 | 3,807 |  |
| 2021 | 3 | 4 | 3 |  | 1,897 | 1,277 |  |
| 2021 | 4 | 1 | 1 |  | 349 | 376 |  |

Table 20.2.4.b. Nephrops FU 20-21. Sampling levels by France.

| FRANCE |  | Number of Samples |  |  | Numbers Measured |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Quarter | Catch | Discards | Landings | Catch | Discards | Landings |
| 2012 | 1 |  | 31 | 9 |  | 391 | 1431 |
| 2012 | 2 |  | 13 | 8 |  | 198 | 1202 |
| 2012 | 3 |  | 47 | 8 |  | 667 | 1155 |
| 2012 | 4 |  | 6 | 6 |  | 16 | 860 |
| 2013 | 1 |  | 0 | 12 |  | 0 | 1362 |
| 2013 | 2 |  | 68 | 72 |  | 1,120 | 3151 |
| 2013 | 3 |  | 16 | 68 |  | 131 | 1917 |
| 2013 | 4 |  | 2 | 14 |  | 12 | 1303 |
| 2014 | 1 |  | 0 | 10 |  | 0 | 1221 |
| 2014 | 2 |  | 40 | 47 |  | 1,127 | 3536 |
| 2014 | 3 |  | 20 | 33 |  | 458 | 1934 |
| 2014 | 4 |  | 0 | 9 |  | 0 | 1360 |
| 2015 | 1 |  | 2 | 14 |  | 60 | 1508 |
| 2015 | 2 |  | 24 | 44 |  | 520 | 3249 |
| 2015 | 3 |  | 1 | 9 |  | 1 | 1366 |
| 2015 | 4 |  | 0 | 9 |  | 0 | 1357 |
| 2016 | 1 |  | 3 | 44 |  | 464 | 3164 |
| 2016 | 2 |  | 4 | 42 |  | 519 | 1263 |
| 2016 | 3 |  | 1 | 25 |  | 217 | 1971 |
| 2016 | 4 |  | 2 | 20 |  | 5 | 1935 |
| 2017 | 1 |  | 3 | 46 |  | 429 | 1659 |
| 2017 | 2 |  | 3 | 80 |  | 852 | 2390 |
| 2017 | 3 |  | 2 | 9 |  | 84 | 344 |
| 2017 | 4 |  | 1 | 23 |  | 307 | 952 |
| 2018 | 1 |  | 8 | 8 |  | 460 | 36 |
| 2018 | 2 |  | 9 | 9 |  | 1190 | 254 |
| 2018 | 3 |  | 30 | 30 |  | 1140 | 105 |
| 2018 | 4 |  | 10 | 10 |  | 149 | 19 |


| FRANCE |  | Number of Samples |  | Numbers Measured |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Year | Quarter | Catch | Discards | Landings | Catch |
| 2019 | 1 | 8 | Discards | Landings |  |
| 2019 | 2 | 9 | 12 | 588 | 51 |
| 2019 | 3 | 30 | 21 | 1,501 | 46 |
| 2019 | 4 | 10 | 5 | 486 | 32 |
| $2020^{*}$ | all | na | na | na | na |
| $2021^{*}$ | all | na | na | na | na |
| $2022^{*}$ | all | na | na | na | na |

*No sampling due to low level fishery participation.

Table 20.2.5. Nephrops FU 20-21. Sex ratio in the landings by country based on available sampling.

| Year | Ireland |  |  |
| :---: | :---: | :---: | :---: |
|  | Females ('000s) | Males ('000s) | \% Males in Landings |
| 2012 | 1,171 | 25,304 | 96 |
| 2013 | 8,369 | 15,596 | 65 |
| 2014 | 13,650 | 25,503 | 65 |
| 2015 | 8,930 | 39,078 | 81 |
| 2016 | 15,807 | 23,835 | 60 |
| 2017 | 11,836 | 29,183 | 71 |
| 2018 | 15,967 | 28,486 | 64 |
| 2019 | 23,578 | 51,264 | 68 |
| 2020 | 2,768 | 9,124 | 77 |
| 2021 | 4,539 | 12,770 | 74 |


| France |  |  |  |
| :---: | :---: | :---: | :---: |
| Year | Females ('000s) | Males ('000s) | \% Males in Landings |
| 2012 | 1,545 | 9,323 | 86 |
| 2013 | 1,678 | 7,641 | 82 |
| 2014 | 3,292 | 7,316 | 69 |
| 2015 | 1,144 | 6,244 | 85 |
| 2016 | 819 | 8,815 | 91 |
| 2017 | 1,119 | 5,110 | 82 |
| 2018 | 1,863 | 3,605 | 66 |
| 2019* | - | - | - |
| 2020** | - | - | - |
| 2021** | - | - | - |

*Sampling data provided but not used due to quality issues.
${ }^{* *}$ No sampling due to low level fishery participation.

Table 20.2.6. Nephrops FU 20-21. Landings and discards by number and weight ( $\mathbf{t}$ ), dead discard rate and discard rate by number, discard rate by weight and estimated mean weights (grs) in the landings and discards for France. 25\% discards survival.

| France |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \grave{亠} \\ & \stackrel{\text { ® }}{2} \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |
|  | millions | millions | millions | \% | \% | \% | tonnes | tonnes | gramme | gramme |
| 2012 | 10.9 | 17.8 | 24.2 | 55.1 | 62.1 | 41.5 | 453 | 322 | 41.7 | 18.1 |
| 2013 | 9.3 | 10.0 | 16.9 | 44.7 | 51.9 | 26.6 | 486 | 176 | 52.2 | 17.6 |
| 2014 | 10.6 | 37.0 | 38.4 | 72.4 | 77.7 | 55.8 | 465 | 588 | 43.8 | 15.9 |
| 2015 | 7.4 | 7.7 | 13.2 | 43.9 | 51.1 | 31.7 | 355 | 165 | 48.1 | 21.4 |
| 2016 | 9.6 | 3.2 | 12.0 | 19.7 | 24.7 | 16.2 | 477 | 92 | 49.5 | 29.1 |
| 2017 | 6.2 | 5.9 | 10.7 | 41.6 | 48.7 | 26.2 | 341 | 121 | 54.8 | 20.5 |
| 2018 | 5.5 | 4.7 | 9.0 | 39.0 | 46.1 | 32.3 | 195 | 93 | 35.6 | 19.9 |
| 2019* | - | - | - | - | - | - | - | - | - | - |
| 2020** | - | - | - | - | - | - | - | - | - | - |
| 2021** | - | - | - | - | - | - | - | - | - | - |

*Sampling data provided but not used due to quality issues.
**Sampling data not available due to low level fishery participation.

Table 20．2．7．Nephrops FU 20－21．Landings and discards by number and weight（ t ），dead discard rate and discard rate by number，discard rate by weight and estimated mean weights（grs）in the landings and discards for Ireland．25\％discards survival．

| Ireland |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { 㐫 } \\ & \text { 见 } \end{aligned}$ |  |  |  |  |  |  | $\begin{aligned} & \text { 以 } \\ & \text { 듬 } \\ & \text { त्0 } \end{aligned}$ |  |  |  |
|  | millions | millions | millions | \％ | \％ | \％ | tonnes | tonnes | gramme | gramme |
| 2012 | 26.5 | 17.5 | 39.6 | 33.1 | 39.7 | 22.6 | 708 | 207 | 26.7 | 11.9 |
| 2013 | 24.2 | 8.3 | 30.5 | 20.5 | 25.6 | 14.0 | 844 | 137 | 34.9 | 16.4 |
| 2014 | 39.1 | 17.6 | 52.3 | 25.3 | 31.1 | 14.8 | 1342 | 233 | 34.3 | 13.3 |
| 2015 | 47.9 | 18.6 | 61.9 | 22.5 | 27.9 | 13.3 | 1620 | 248 | 33.8 | 13.4 |
| 2016 | 39.6 | 27.5 | 60.3 | 34.2 | 41.0 | 26.9 | 1531 | 564 | 38.6 | 20.5 |
| 2017 | 41.0 | 9.2 | 47.9 | 14.4 | 18.4 | 9.7 | 1113 | 120 | 27.1 | 13.0 |
| 2018 | 44.5 | 11.9 | 53.4 | 16.8 | 21.2 | 14.4 | 1197 | 201 | 26.9 | 16.9 |
| 2019＊ | 74.8 | 29.2 | 96.7 | 22.6 | 28.1 | 16.5 | 2219 | 439 | 29.7 | 15.0 |
| 2020 | 11.9 | 1.7 | 13.1 | 9.5 | 12.3 | 7.6 | 336 | 28 | 28.2 | 16.7 |
| 2021 | 17.3 | 2.3 | 19.0 | 8.9 | 11.5 | 6.3 | 572 | 38 | 33.0 | 17.0 |

＊2019 data revision due to valid sample inclusion．

Table 20.2.8. Nephrops FU 20-21. Landings and discards by number and weight ( $\mathbf{t}$ ), dead discard rate and discard rate by number, discard rate by weight and estimated mean weights (grs) in the landings combined by both countries based on available sampling and scaled to international landings. 25\% discards survival.

| Combined and scaled to the international landings |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \frac{\grave{\pi}}{\pi} \\ & \text { ঠ} \end{aligned}$ |  |  |  |  |  |  |  |  |  | Mean weight in discards |
|  | millions | millions | millions | \% | \% | \% | tonnes | tonnes | gramme | gramme |
| 2012 | 38.2 | 36.1 | 65.3 | 41.4 | 48.5 | 31.3 | 1,189 | 542 | 31.1 | 15.0 |
| 2013 | 34.8 | 19.2 | 49.2 | 29.3 | 35.6 | 19.1 | 1,387 | 327 | 39.9 | 17.0 |
| 2014 | 50.6 | 55.5 | 92.2 | 45.2 | 52.3 | 31.2 | 1,836 | 834 | 36.3 | 15.0 |
| 2015 | 59.4 | 28.1 | 80.5 | 26.2 | 32.2 | 17.3 | 2,116 | 442 | 35.7 | 15.7 |
| 2016 | 60.2 | 37.5 | 88.3 | 31.8 | 38.4 | 24.6 | 2,453 | 801 | 40.7 | 21.4 |
| 2017 | 60.1 | 19.2 | 74.5 | 19.4 | 24.3 | 14.2 | 1,849 | 306 | 30.8 | 15.9 |
| 2018 | 64.7 | 21.5 | 80.8 | 20.0 | 25.0 | 17.5 | 1,803 | 381 | 27.9 | 17.7 |
| 2019* | 91.8 | 35.8 | 118.7 | 22.6 | 28.1 | 16.5 | 2,723 | 539 | 29.7 | 15.0 |
| 2020 | 14.6 | 2.0 | 16.2 | 9.5 | 12.3 | 7.6 | 413 | 34 | 28.2 | 16.7 |
| 2021 | 22.3 | 2.9 | 24.4 | 8.9 | 11.5 | 6.3 | 736 | 49 | 33.0 | 17.0 |

*2019 data revision due to valid sample inclusion for Ireland and revision to UK landings.

Table 20.2.9. Nephrops FU 20-21. Results summary table for geo-statistical analysis of UWTV survey.

| Ground | Year | Number of stations | Mean Density** (burrows/m²) | Domain <br> Area <br> ( $\mathrm{Km}^{2}$ ) | Geostatistical Abundance <br> Estimate adjusted (millions burrows) | CV on <br> Burrow estimate <br> (\%) | Analysis Method software |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FU 2021 | 2012 | 54 | 0.57 |  | $n \mathrm{r}$ | $n \mathrm{r}$ | na |
|  | 2013* | 55 | 0.16 | 10, 014 | 1640 | 8.1 | RGeostats |
|  | 2014 | 98 | 0.19 | 10, 014 | 2021 | 3.9 | RGeostats |
|  | 2015 | 96 | 0.2 | 10,014 | 2003 | 3.2 | RGeostats |
|  | 2016 | 93 | 0.18 | 10,014 | 1879 | 4.3 | RGeostats |
|  | 2017 | 86 | 0.44 | 10,014 | 4428 | 3.8 | RGeostats |
|  | 2018 | 96 | 0.27 | 10,014 | 2721 | 4.0 | RGeostats |
|  | 2019 | 95 | 0.06 | 10,014 | 617 | 4.8 | RGeostats |
|  | 2020 | 97 | 0.10 | 10,014 | 1020 | 4.8 | RGeostats |
|  | 2021 | 97 | 0.12 | 10, 014 | 1202 | 3.9 | RGeostats |
|  | 2022 | 92 | 0.10 | 10, 014 | 1032 | 4.8 | RGeostats |

[^13]Table 20.3.1. Nephrops FU 20-21. Short-term catch options prediction inputs and recent estimates of mean weight in landings and harvest rates. Cells in bold indicates inputs to catch option calculations.

|  |  |  |  |  | Discard Proportion number |  |  |  |  |  | Mean weight in landings |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \frac{1}{\pi} \\ & \stackrel{\text { ® }}{2} \end{aligned}$ | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
| 2012 | 38.2 | 36.1 | 65.3 | 41.4 | 48.5 |  |  |  | 1189 | 542 | 31.1 | 15.0 |
| 2013 | 34.8 | 19.2 | 49.2 | 29.3 | 35.6 | 1640 | 261 | 3.0 | 1387 | 327 | 39.9 | 17.0 |
| 2014 | 50.6 | 55.5 | 92.2 | 45.2 | 52.3 | 2021 | 154 | 4.6 | 1836 | 834 | 36.3 | 15.0 |
| 2015 | 59.4 | 28.1 | 80.5 | 26.2 | 32.2 | 2003 | 129 | 4.0 | 2116 | 442 | 35.7 | 15.7 |
| 2016 | 60.2 | 37.5 | 88.3 | 31.8 | 38.4 | 1879 | 157 | 4.7 | 2453 | 801 | 40.7 | 21.4 |
| 2017 | 60.1 | 19.2 | 74.5 | 19.4 | 24.3 | 4428 | 332 | 1.7 | 1849 | 306 | 30.8 | 15.9 |
| 2018 | 64.7 | 21.5 | 80.8 | 20.0 | 25.0 | 2721 | 212 | 3.0 | 1803 | 381 | 27.9 | 17.7 |
| 2019 | 91.8 | 35.8 | 118.7 | 22.6 | 28.1 | 617 | 58 | 19.2 | 2723 | 539 | 29.7 | 15.0 |
| 2020 | 14.6 | 2.0 | 16.2 | 9.5 | 12.3 | 1020 | 96 | 1.6 | 413 | 34 | 28.2 | 16.7 |
| 2021 | 22.3 | 2.9 | 24.2 | 8.9 | 11.5 | 1202 | 92 | 2.0 | 736 | 49 | 33.0 | 17.0 |
|  |  |  |  | 13.7 | 17.3 |  |  |  |  |  | 30.3 | 16.2 |

Average 2019-2021


Figure 20.1.1. Nephrops FU 20-21. Number of Irish vessels reporting landings $\mathbf{> 1 0} \mathbf{t}$ by year.


Figure 20.1.2. Nephrops FU 20-21. Combined box and kite plot of vessel power on the FU20-21 grounds by year. The blue line indicates the mean.

met.short
FPO_CRU_0_0_0_all
MIS_MIS_O_0_0_all OTB_CRU_100-119 OTB_CRU_70-99 OTB_DEF_100-119
OTB DEF_70-99
SSC_DEF_100-119_0_0_a
TBB_DEF_100-119_0_0_a
TBB_DEF_70-99_0_0_all

Figure 20.1.3. Nephrops FU 20-21. Irish Landings by DCF Metiér.


Figure 20.2.1. Nephrops FU 20-21. Landings in tonnes by country.


Figure 20.2.2. Nephrops FU 20-21. Effort data (Kw days) for the Irish otter trawl Nephrops directed fleet.


Figure 20.2.3. Nephrops FU 20-21. Commercial length-frequency distribution by country. Minimum conservation reference size of 25 CL mm (European MCR) and 35 CL mm (French MLS) displayed. 2019 data provided by France but not included in the assessment. Data not available for France since 2020 due to low fishery participation.


Figure 20.2.4. Nephrops FU 20-21. Annual mean weights (gr) in the landings (blue line) and discards (red line) by country and combined scaled to international landings.


Figure 20.2.5. Nephrops FU 20-21. Annual discard ogive derived from Irish sampling. Minimum landing size of $\mathbf{2 5} \mathbf{C L} \mathbf{~ m m}$ (European MCR) as black line.


Figure 20.2.6. Nephrops FU 20-21. Contour plots of krigged density estimates for the UWTV surveys from 2013 to 2022.


Figure 20.2.8. Nephrops FU 20-21. Time-series of abundance estimates (millions burrows) for FU20-21 (error bars indicate $95 \%$ confidence intervals) and MSY $\mathrm{B}_{\text {trigger }}$ is dashed line.


Figure 20.2.9. Nephrops FU 20-21. Mean size trends for catches by sex from the IGFS- WIBTS-Q4 [G7212] Irish survey. Vertical lines displayed are Minimum Conservation Reference Size $\mathbf{2 5} \mathbf{~ m m}$ Carapace Length (CL) and 35 mm CL.


Figure 20.2.10. Nephrops FU 20-21. Mean size trends for catches by sex from the EVHOE- WIBTS-Q4 [G9527] French survey. No survey data available for 2017. Vertical lines displayed are Minimum Conservation Reference Size 25 mm Carapace Length (CL) and 35 mm CL.


Figure 20.3.11. Nephrops FU 20-21. Harvest rate (\% dead removed / UWTV abundance). The dashed and solid lines are the MSY proxy and the harvest rate respectively.

# 21 Norway lobster (Nephrops norvegicus) in divisions 7.g and 7.f, Functional Unit 22 (Celtic Sea, Bristol Channel) 

## Type of assessment in 2022

UWTV based assessment using WKNEPH 2009 protocol as described in the stock annex. The TV survey is due to be repeated in the summer 2022and the new survey will form the basis of advice for this stock in the autumn. This stock assessment is available in the ICES Transparent Assessment Framework (TAF) here.

## ICES advice applicable to 2021

"ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, catches in 2021 that correspond to the $F$ ranges in the MAP are between 1238 tonnes and 1560 tonnes, assuming recent discard rates. The entire range is considered precautionary when applying the ICES advice rule.
To ensure that the stock in Functional Unit (FU) 22 is exploited sustainably, management should be implemented at the functional unit level."

## ICES advice applicable to 2022

"ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2018-2020, catches in 2022 should be no more than 1257 tonnes.
To ensure that the stock in Functional Unit (FU) 22 is exploited sustainably, management should be implemented at the FU level.
ICES notes the existence of a management plan, developed and adopted by one of the relevant management authorities for Subarea 7. ICES considers this plan to be precautionary when implemented at the FU level."

### 21.1 General

## Stock description and management units

The Smalls Nephrops stock (FU22) covers ICES rectangles 31-32E3, 31-32E4 within 7.f.g. It is included in the whole ICES Area 7 together with Irish Sea East and West [FU14, FU15], Porcupine Bank [FU16], Aran Grounds [FU17], northwest Irish Coast [FU18], southeast and southwest Irish Coast [FU19], NW Labadie, Baltimore and Galley [FU20-21], Jones and Cockburn [FU21].
Historically FU20-22 has covered an amalgamation of several spatially distinct mud patches; FU 20 NW Labadie, Baltimore and Galley, FU 21 Jones and Cockburn and FU22 the Smalls. There is no evidence that the whole exploited area belongs to the same stock or that there are several
patches linked in meta-population sense. WGCSE 2013 recommended that FU20-22 should be split into FU20-21 and FU22 for the purposes of assessment and advice provision. The map below shows FU22 assessment area (blue) and TAC area (red). There is no evidence that the individual functional units belong to the same stock. See Section 18 for details on Nephrops in Subarea 7 general section.


## Ecosystem aspects

This section is detailed in stock annex.

## Fishery description

Ireland, France and the UK are the main countries involved in the FU22 Nephrops fishery. In the early 2000 s the Irish fleet had on average over $70 \%$ of the landings and this has increased to over $85 \%$ from this FU in recent times. A description of this fleet is given in the stock annex. The time-series of numbers of vessels is updated in Figure 21.1.1. The numbers of vessels has been decreasing in recent years where the highest number was recorded in 2016. The time-series of vessel power is shown as a box and kite plot in Figure 21.1.2.

Irish landings from this FU come mainly from ICES statistical rectangle 31E3. The fishery on the Smalls grounds operates throughout the year, weather permitting with a seasonal trend.

French trawlers targeting Nephrops in the Celtic Sea operate mainly in FU20-21. In the early 2000s French fleet had on average $30 \%$ of the landings from FU22 where this has decreased to $<1 \%$ in recent times. 80-90\% of the FU22 French landings come from ICES statistical rectangle 31E3.

UK fleet had on average $\sim 10 \%$ of the landings in recent year and is mainly UK-Northern Irish vessels in this fishery.

## Fishery in 2021

In 2021, 60 Irish vessels reported landings from FU22. Of these, 52 vessels reported landings in excess of 10 t . Vessels $>18 \mathrm{~m}$ account for $90 \%$ of the landings in 2021. In recent years several newer vessels specializing in Nephrops fishing have participated periodically in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates.

In 2021, ten French trawlers reported landings for FU22. French vessels switch between FU2021 and FU22. In 2021, one Northern Ireland and UK-Scotland and five UK(E\&W) vessels reported landings for this FU.

The French minimum mesh size of codend was set at 100 mm since January 2000 the majority of Irish landings are from vessels with $80-99 \mathrm{~mm}$ codend mesh.

## Information from stakeholders

None presented.

### 21.2 Data

## InterCatch

Data were available in InterCatch and used for catch data only. French catch data provided directly by the national expert.

## Landings

There was a 3\% increase in 2020 landings reported by Ireland which affected quarter 1 and 3 only. The sample data was revised using this updated landings data and used in the assessment this year.

The reported landings time-series by country is shown in Figure 21.2.1 and Table 21.2.1. The reported Irish landings from FU22 have increased since 2000. In 2020 the landings increased from 2019 by $23 \%$ to approximately 1448 t . French landings have gradually decreased since the early 2000s to the present. Reported landings from the UK have fluctuated with a decrease in 2020. Northern Ireland had the highest landings at 16 t followed by England and Wales reporting 6 t . Belgium reported minimal landings $<2.5 \mathrm{t}$ in general from this FU.

## Effort

In line with WGCSE 2015 recommendation effort is reported in Kwdays and lpue reported in $\mathrm{t} / \mathrm{Kwdays}$ in the knowledge that the trend is likely to be a biased underestimate because it is not adjusted for efficiency or behavioural changes. The effort series is based on the same criteria for FU15, 16, 17, 22 and 20-21 ( $30 \%$ landings threshold) and will be contingent on the accuracy of landings data reported in logbooks. Effort data are available for the Irish Nephrops directed fleet in FU22 from 1995-2021. The time-series of effort and lpue is updated in Figure 21.2.2 and Table 21.2.2.

Effort shows an increasing trend since the early 2000s (Table 21.2.2. and Figure 21.2.2) with a decreasing trend since 2018.

## Sampling levels

Dedicated sampling of landings and discards began in 2003 by Ireland. Sampling levels in 2021 were good (Figure 21.2.3).

## Sampling and Raising Procedure Review

The national sample raising procedures for FU22 were reviewed and fully documented through an R markdown document (Annex 3, ICES, 2018 and stock annex). Annual discard ogives are
calculated and are applied to quarterly length distributions and then raised to total quarterly landings before aggregation. A further raising procedure is applied to raise the annual sampled Irish data, where this addresses quarters with missing length samples. Next the international raising factor is applied. This raising procedure is used to assess this stock and to calculate mean weights, sex ratio and discard rates as inputs for catch scenarios and advice. A minor data revision to 2018 sample data was presented to WGCSE 2020 and resulting calculations were accepted. The revision to 2020 landings data for Ireland resulted in an increase in numbers in landings, discards and removals. This resulted in an increase in the harvest rate for 2020 to $10.1 \%$ (previous 9.7\%) which was still below FmsY.

## Commercial length-frequency distributions

The Irish sampling programme started in 2003 and since then coverage and intensity have been very good covering the seasonal trend of the fishery. The mean size of Nephrops in Irish landings has increased in the recent 3 years for both sexes. The mean size of Nephrops in the catch has remained relatively stable since 2005 (Figure 21.2.5). There is an increase in mean size in the catches in 2007 to 2009 for both sexes which is linked to the recruitment signal picked up by both the 2006 UWTV [U5917] and IGFS-WIBTS-Q4 [G7212].

## Sex ratio

The sex ratio by year is shown in Figure 21.2.6. This shows some fluctuations over time. The sex ratio has a distinct seasonal pattern (Figure 21.2.7) with lowest male proportions in the samples in May and June. Males dominate the catches in the autumn and winter.

## Mean weight explorations

Explorations of the mean weight in the catch samples by sex shows a strong cyclical pattern in the females (Figure 21.2.7). This corresponds with the emergence of mature females from the burrows to mate in summer. There is an increase in mean weight in 2007 to 2009 for both sexes which is linked to the recruitment signal picked up by both the UWTV [U5917] and IGFS-WIBTSQ4 [G7212] (Figure 21.2.11). The annual mean weight estimate for landings and discards is shown in Figure 21.2.8. The mean weight estimates in the landings show an increasing trend since 2019.

## Discarding

Since 2003 discard rates have been estimated using unsorted catch and discards sampling. This involves unsorted catch and discard samples being provided by vessels or collected by observers at sea on discard trips. The catch sample is partitioned into landings and discards using an onboard discard selection ogive derived for the discard samples. Sampling effort is stratified monthly, but annual aggregations are used to derive length distributions and selection ogives. The length-weight regression parameters given in the stock annex are used to calculate sampled weights and appropriate quarterly raising factors. The sampling intensity and coverage has varied over the time-series, but overall has been good.

Discard rates range between $9-39 \%$ of total catch by weight and $15-52 \%$ of total catch by number (Table 21.2.4). Discard rate of females tends to be higher due to the smaller average size and market reasons. There is no information on discard survival rate in this fishery. $25 \%$ is assumed in line with other Nephrops stocks in the Celtic Sea (Charuau et al., 1982). Highest discard rates were observed in 2007 as a result of the recruitment into the fishery in 2006.

Gear selectivity trials by Bord Iascaigh Mhara (BIM, 2017) reported a $64 \%$ survivor rate for Nephrops caught in a trawl with a SELTRA selectivity device in the outer Galway Bay area.

Table 21.3.1 gives weights, numbers and mean weights of the landings and discard raised internationally according to the stock annex.

## Surveys

## Abundance indices from UWTV surveys

The methods used during the survey were similar to those employed for UWTV surveys [U5917] of Nephrops stocks around Ireland and elsewhere are documented by WKNEPHTV (ICES, 2007), WKNEPHBID (ICES, 2008), SGNEPS (ICES, 2009; 2010; 2012), WGNEPS (ICES, 2013; 2014; 2015; 2016a; 2017; 2018a, 2020, 2021, 2022), WKNEPS (ICES, 2016b; 2018b), Leocádio, A., et al, 2018 and Dobby H., et al, 2021.

SGNEPS 2012 (ICES, 2012) recommended that a CV (or relative standard error) of $<20 \%$ is an acceptable precision level for UWTV survey estimates of abundance. This allowed sampling intensity to be reduced from around 90 stations in the past to around 41 on the Smalls grounds in 2022 which allowed survey coverage of other FUs. A randomised isometric grid design was employed with UWTV stations at 4.5 nmi intervals, whereas previously a 3.0 nmi square grid was used. Operational details of the 2022 UWTV survey are available (Doyle et al., 2022).

Seven stations in FU22 were not surveyed successfully in 2015 due to very poor visibility conditions encountered as a result of strong tides. WKCELT 2014 concluded that WGCSE or WGNEPS should make recommendations on the most appropriate fill in procedure to be adopted in cases when stations could not be surveyed. WGCSE 2015 agreed the following procedure for this case: Two buffer zones of 1 nmi and 2 nmi distance were generated around the missing stations. The counts and mean of historic density estimates within the 1 and 2 nmi buffers were calculated. The standard kriging procedure was carried out and summary results were computed for the 1 and 2 nmi "fill-ins". Finally the mean of historic densities within 2 nmi buffer of the planned stations were used in the calculation of the 2015 abundance.

The blanked krigged contour plot and posted point density data are shown in Figure 21.2.9. The krigged contours correspond very well to the observed data. In general, the densities are higher in the central area of the ground with a localised hotspot centrally and also in the southwestern leg. Densities and abundance have remained stable for most of the time-series with the exception of the first year and 2017, which were the highest in the series. The 2022 mean density 0.31 burrows $/ \mathrm{m}^{2}$ is approximately $30 \%$ increase compared with density 0.23 burrows $/ \mathrm{m}^{2}$ in 2021 . The summary statistics from this geostatistical analysis are given in Table 21.2.5 and plotted in Figure 21.2.10. The geostatistical abundance estimate adjusted is derived using the mean of the krigged grid where the mean of the observations is reported in Table 21.2.5.

The 2022 estimate of 895 million burrows is below the MSY Btrigger ( 990 million). The estimation variance of the survey as calculated by EVA is very low (CVs in the order $<9 \%$ ).

## Groundfish survey data

The Irish groundfish survey IGFS-WIBTS-Q4 [G7212] and French EVHOE- WIBTS-Q4 [G9527] survey operate in the Celtic Sea (Stokes et al., 2014; ICES, 2017b). These provide information on length-frequency compositions, mean size in the catches, cpue of Nephrops in FU22. The mean size of the catches is stable over the time-series except in 2006 and 2008, which signals recruitment into the fishery in 2006 and 2007 (Figure 21.2.11 and 21.2.12). This signal of recruitment was also picked up during the 2006 UWTV [U5917] survey (Doyle et al., 2012). The groundfish surveys in the Celtic Sea provide a useful indicator of recruitment in this FU.

### 21.3 Assessment

## Comparison with previous assessments

The WGCSE 2021 carried out an UWTV-based assessment for this stock. The methods used were very much in line with WKNEPH (ICES, 2009) and the approach taken for other Nephrops stocks in areas 6 and 7 by WGCSE.

## State of the stock

UWTV abundance estimates suggest that the stock size shows a recent declining trend with a an increase in 2022. The 2022 estimate is below the MSY $B_{\text {trigger }}$ ( 990 million). The 2022 estimate ( 895 million) is below the average of the series (geomean [2006-2022]: 1125 million).

Harvest rate is calculated as (landings + dead discards)/(abundance estimate). Table 21.3.1 and Figure 21.3.1 summarize recent harvest rates. Recent harvest rates have fluctuated due to recruitment pulses into the fishery in 2006 and 2010 and is currently $10.7 \%$ which is below $\mathrm{F}_{\text {msy. }}$.

### 21.4 Catch scenarios table

Catch scenario table inputs and historical estimates of mean weight in landings and harvest rates are presented in Table 21.3.1 and summarised below.

Since 2003, mean weight in the landings has varied between $18-27$ grammes (Figure 21.2.8). WGCSE 2019 decided that given the stability in mean weights in the recent years, the recent three year average of mean weights is to be used to calculate catch scenarios. The three year average (2019-2021) of proportion of removals retained was used as is standard for other Nephrops stocks. The estimate harvest rate has also varied a lot, from 6-27\% with 2007 being the highest observed (Figure 21.3.1). This is a result of recruitment into the fishery in 2006 and 2007.
The basis for the catch scenarios:

| Variable | Value | Notes |
| :--- | :---: | :--- |
| Stock abundance (2023) | 895 | Number of individuals (million); UWTV survey 2022 |
| Mean weight in projected landings | 25.9 | Average 2019-2021 in grammes |
| Mean weight in projected discards | 13.4 | Average 2019-2021 in grammes |
| Projected discards | 20.4 | Proportion by number; average 2019-2021 |
| Discards survival | 25 | Proportion by number |
| Projected dead discards | 16.2 | Proportion by number; average 2019-2021 |

A prediction of landings for FU22 using the approach agreed procedure proposed at WKNEPH 2009 and outlined in the stock annex will be made on the basis of the 2022 UWTV survey. This will be presented in October 2022 for the provision of advice.

### 21.5 Reference points

New reference points were derived by WKMSYRef4 (ICES, 2016XX, 2016YY) for FU22. These were updated on the basis of an average of estimated Fmsy proxy harvest rates over a period of years, this corresponds more closely to the methodology for finfish. In cases where there is a clear trend in the values a five year average was chosen. Similarly, the five year average of the F at $95 \%$ of the YPR obtained at the FMSY proxy reference point was proposed as the Fmsy lower bound and the five year average of the F above $\mathrm{F}_{\max }$ that leads to YPR of $95 \%$ of the maximum as the upper bound. Using an average value also has the advantage of reducing the effect of any unusually high or low estimates of the FMSY proxy which occasionally appear.

This stock previously did not have MSY $B_{\text {trigger }}$ specified, the time-series and range of indicator biomass is also limited such that direct use of Bloss is considered too close to equilibrium biomass. The workshop proposed to use the $5 \%$ interval on the probability distribution of indicator biomass assuming a normal distribution, which is analogous to the $5 \%$ on Bmsy proposed for finfish stocks assuming these Nephrops FU's have been exploited at a rate close to near HRmsy. The MSY $B_{\text {trigger }}$ for FU22 is 987 million individuals rounded to 990 million.

| Stock code | MSY Flower* | F MSY* $^{*}$ | MSY Fupper* with AR | MSY B Brigger | MSY Fupper* with no AR |
| :--- | :---: | :---: | :---: | :---: | :---: |
| nep-22 | $10.2 \%$ | $12.8 \%$ | $12.8 \%$ | $990^{* * *}$ | $12.8 \%$ |

* Harvest rate (HR).
*** Abundance in millions.


### 21.6 Management strategies

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to Norway lobster (Nephrops norvegicus) by functional unit in ICES subarea 7 and also demersal stocks. There is currently no agreement with the UK regarding this plan.

### 21.7 Quality of assessment and forecast

Since 2006, a dedicated annual UWTV survey has provided abundance estimates for FU22 with high precision. There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008; SGNEPS 2009; WGNEPS 2016, WGNEPS 2018b). Ultimately there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that is more accurate, although no more precise (WKNEPH 2009). The survey estimates themselves are very precisely estimated (CVs 2-9\%) given the homogeneous distribution of burrow density and the modelling of spatial structuring. The cumulative bias estimates for FU22 are largely based on expert opinion. The precision of these bias corrections cannot yet be characterised, but is likely to be lower than that observed in the survey.

In 2015, there is added uncertainty, not accounted for in the model or CV estimate, because $17 \%$ of the planned TV stations could not be successfully surveyed due to poor visibility on the seabed. However, the spatial distributions of densities have been fairly consistent over time and the overall density has also been relatively stable. The fill in procedure used to generate density estimates for the seven missing stations should be a good approximation.

A review of sampling and raising procedures was presented to WGCSE 2018 and is accepted as the current method to calculate the fishery dependant inputs FU22 (Annex 3, ICES, 2018 and stock annex).

In the provision of catch scenarios based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. Given the recent stability in mean weights in landings and unwanted catch - for FU22 deterministic estimates of the mean weight in the landings and discard rates for 2019-2021 are used by the WG. Previously the full time series was used to account for the variability over time where this had occurred when large recruitments are observed in the stock as was the case in 2006 and 2007.

From 2016, fisheries catching Nephrops in Subarea 7 are covered by the EU landings obligation with several exemptions (EU, 2015). The average discard rate by weight for FU22 over the last three years is $12 \%$. Irish discard survival experiments indicate that the trawl discard survival may be around $64 \%$ (BIM, 2017). As a result, an exemption from the landings obligation based on high survivability has been granted by the European Commission. Catch advice and scenarios are provided this year on the assumption that discarding is assumed to continue at the recent average.

Landings data are adjusted to take into account landings that have been misreported from FU16 since 2011. This adjustment is thought to be reasonably accurate (See Section 20).

Sampling and discard estimates have improved over the time-series.

### 21.8 Recommendation for next benchmark

This stock has not been formally benchmarked by ICES although the approach used has. WGCSE recommends that the issue list below can be addressed through an inter-bench process:

- The biological parameters used as inputs to the SCA should be reconsidered; growth parameters, length-at-maturity and natural mortality.
- The historical time-series of landings and effort by rectangle should be disaggregated and options for standardisation of lpue investigated.
- Historical sampling and groundfish survey data in this FU should also be disaggregated as far as possible back in time, and investigated for useful trends and signals.


### 21.9 Management considerations

The trends from the fishery (landings, effort, mean size, etc.) appear to show stock is exploited. The UWTV abundance and mean density estimates show some fluctuations in burrow abundance in the recent years. There are fluctuations in the harvest rates which are related to the signals of recruitment into the fishery in 2006 and 2007 picked up by the UWTV survey and IGFS-WIBTS-Q4. Recent harvest rates for the FU22 Smalls fluctuate and suggest the stock is exploited below FMSY.

A new survey point available in September 2022 will provide a more up to date prognosis of stock status. This up to date survey information will be used to generate catch options and the provision of advice in October 2022.

In recent years, several newer vessels specializing in Nephrops fishing have participated in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimize catch rates. There have been concerns that effort could be displaced towards the Smalls and other Nephrops grounds due to effort controls in 7.a and 6.a. This has not happened to date, and the 2014 and 2020 effort was just below the recent average in the time-series.

There has been a trend for Irish vessels ( $>18 \mathrm{~m}$ ) to switch to multi (quad) rig trawls. Provisional data suggest a $\sim 30 \%$ increase in Nephrops catch rates and a reduction in fish bycatch of $\sim 30 \%$ due to the lower headline height.

Nephrops fisheries in the Smalls have non-Nephrops bycatch composition. Cod, whiting and to a lesser extent haddock are the main bycatch species (Davie and Lordan, 2011). A targetted whiting fishery also overlaps with the Nephrops fishery in this area, but this has negligible bycatch of Nephrops.

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Table 21.2.1. Nephrops in FU22 (Smalls Grounds). Landings in tonnes by country.

| FU 22 Landings (t) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | France | Rep. of Ireland | UK | Belgium | Total |
| 1999 | 1034 | 741 | 0 |  | 1775 |
| 2000 | 1192 | 1687 | 11 |  | 2890 |
| 2001 | 882 | 2054 | 2 |  | 2938 |
| 2002 | 598 | 1392 | 3 |  | 1993 |
| 2003 | 799 | 1257 | 10 |  | 2065 |
| 2004 | 454 | 1349 | 26 |  | 1828 |
| 2005 | 478 | 1987 | 68 |  | 2533 |
| 2006 | 293 | 1442 | 19 | 7 | 1761 |
| 2007 | 216 | 2716 | 13 | 5 | 2950 |
| 2008 | 301 | 2539 | 241 | 9 | 3090 |
| 2009 | 258 | 1609 | 306 | 12 | 2185 |
| 2010 | 129 | 2219 | 351 | 15 | 2714 |
| 2011 | 64 | 1521 | 44 | 7 | 1636 |
| 2012 | 65 | 2506 | 41 | 6 | 2618 |
| 2013 | 83 | 2054 | 107 | 12 | 2257 |
| 2014 | 29 | 2428 | 61 | 8 | 2526 |
| 2015 | 9 | 2215 | 121 | 5 | 2350 |
| 2016 | 5 | 2967 | 354 | 3 | 3329 |
| 2017 | 7 | 2815 | 737 | 1 | 3560 |
| 2018 | 3 | 1639 | 331 | 1 | 1974 |
| 2019 | 9 | 1884 | 187 | 2 | 2083 |
| 2020 | 3 | 1491 | 22 | 2 | 1518 |
| 2021 | <1 | 1537 | 69 | 10 | 1611 |

Table 21.2.2. Nephrops in FU22 (Smalls Grounds). Effort data for the Irish otter trawl Nephrops directed fleet.

| Year | Effort ( ${ }^{\text {O }}$ ( ${ }^{\text {as Kw Days) }}$ | Landings (tonnes) | Ipue (t/KwDays) |
| :---: | :---: | :---: | :---: |
| 1995 | 552 | 1226 | 2.2 |
| 1996 | 412 | 1010 | 2.5 |
| 1997 | 474 | 1096 | 2.3 |
| 1998 | 524 | 1353 | 2.6 |
| 1999 | 292 | 620 | 2.1 |
| 2000 | 586 | 1335 | 2.3 |
| 2001 | 789 | 1964 | 2.5 |
| 2002 | 615 | 1298 | 2.1 |
| 2003 | 639 | 1000 | 1.6 |
| 2004 | 620 | 981 | 1.6 |
| 2005 | 986 | 1882 | 1.9 |
| 2006 | 855 | 1374 | 1.6 |
| 2007 | 1131 | 2677 | 2.4 |
| 2008 | 1047 | 2501 | 2.4 |
| 2009 | 702 | 1605 | 2.3 |
| 2010 | 962 | 2198 | 2.3 |
| 2011 | 724 | 1497 | 2.1 |
| 2012 | 970 | 2260 | 2.3 |
| 2013 | 902 | 1849 | 2.0 |
| 2014 | 915 | 2182 | 2.4 |
| 2015 | 971 | 2076 | 2.1 |
| 2016 | 1270 | 2761 | 2.2 |
| 2017 | 1229 | 2712 | 2.2 |
| 2018 | 748 | 1509 | 2.0 |
| 2019 | 786 | 1736 | 2.2 |
| 2020 | 681 | 1408 | 2.1 |
| 2021 | 666 | 1450 | 2.2 |


|  | Table 21.2.4. Nephrops in FU22 (Smalls Grounds). Landings and discards weight and numbers by year from Irish sampling |  |
| :--- | :--- | :--- | :--- | :--- |
| programme. |  |  |

Table 21.2.5. Nephrops in FU22 (Smalls Grounds). Results summary table for geostatistical analysis of UWTV survey.

| Year | Number of stations | Mean Density <br> adjusted** <br> (burrows $/ \mathrm{m}^{2}$ ) | Domain Area ( $\mathrm{km}^{2}$ ) | Geostatistical <br> Abundance Estimate <br> adjusted <br> (millions burrows) | CV on <br> Burrow estimate \% |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 100 | 0.49 | 2962 | 1503 | 2.4 |
| 2007 | 107 | 0.37 | 2955 | 1136 | 5.7 |
| 2008 | 76 | 0.36 | 2698 | 1114 | 5.6 |
| 2009 | 67 | 0.36 | 2824 | 1093 | 5.0 |
| 2010 | 90 | 0.37 | 2861 | 1141 | 3.9 |
| 2011 | 107 | 0.41 | 2881 | 1256 | 2.9 |
| 2012* | 47 | 0.49 | 2934 | 1498 | 8.1 |
| 2013* | 41 | 0.41 | 2975 | 1254 | 7.2 |
| 2014* | 52 | 0.53 | 2970 | 1622 | 8.4 |
| 2015* | 40 | 0.49 | 3064 | 1363 | 7.0 |
| 2016* | 41 | 0.31 | 3063 | 866 | 6.6 |
| 2017* | 40 | 0.55 | 3063 | 1600 | 4.9 |
| 2018* | 42 | 0.31 | 3063 | 876 | 9.0 |
| 2019* | 41 | 0.40 | 3063 | 1121 | 6.4 |
| 2020* | 40 | 0.27 | 3063 | 750 | 8.0 |
| 2021* | 42 | 0.23 | 3063 | 656 | 6.7 |
| 2022* | 41 | 0.31 | 3063 | 895 | 6.5 |

* reduced isometric grid 4.5 nmi
** mean density adjusted of the observations.

Table 21.3.1. Nephrops in FU22 (Smalls Grounds). Short-term catch option prediction inputs and recent estimates of mean weight in landings and harvest rate (cells in bold indicates inputs to catch scenario calculations).

|  |  |  |  |  |  | UWTV abundance estimate |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
| 2003 | 95.2 | 67.6 | 145.8 | 34.7 | 41.5 | NA | NA | NA | 2,065 | 720 | 21.7 | 10.7 |
| 2004 | 70.7 | 13.1 | 80.5 | 12.2 | 15.6 | NA | NA | NA | 1,828 | 202 | 25.9 | 15.4 |
| 2005 | 119.3 | 128.6 | 215.7 | 44.7 | 51.9 | NA | NA | NA | 2,533 | 1648 | 21.2 | 12.8 |
| 2006 | 100.2 | 45.2 | 134.1 | 25.3 | 31.1 | 1503 | 70 | 8.9 | 1,761 | 454 | 17.6 | 10.1 |
| 2007 | 165.2 | 180.9 | 300.8 | 45.1 | 52.3 | 1136 | 126 | 26.5 | 2,950 | 1906 | 17.9 | 10.5 |
| 2008 | 143.6 | 26.0 | 163.1 | 12.0 | 15.3 | 1114 | 123 | 14.6 | 3,090 | 289 | 21.5 | 11.1 |
| 2009 | 92.0 | 33.0 | 116.8 | 21.2 | 26.4 | 1093 | 108 | 10.7 | 2,185 | 371 | 23.7 | 11.3 |
| 2010 | 121.8 | 44.5 | 155.2 | 21.5 | 26.8 | 1141 | 88 | 13.6 | 2,714 | 636 | 22.3 | 14.3 |
| 2011 | 60.0 | 13.2 | 69.8 | 14.1 | 18.0 | 1256 | 72 | 5.6 | 1,636 | 196 | 27.3 | 14.9 |
| 2012 | 120.3 | 31.4 | 143.9 | 16.3 | 20.7 | 1498 | 239 | 9.6 | 2,618 | 347 | 21.8 | 11.1 |
| 2013 | 93.5 | 40.1 | 123.6 | 24.3 | 30.0 | 1254 | 177 | 9.9 | 2,257 | 497 | 24.1 | 12.4 |
| 2014 | 100.2 | 33.4 | 125.2 | 20.0 | 25.0 | 1622 | 268 | 7.7 | 2,526 | 460 | 25.2 | 13.8 |


|  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \grave{\pi} \\ & \stackrel{y}{*} \end{aligned}$ | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
|  |  |  |  |  |  | UWTV abundance estimate |  |  |  |  |  |  |
| $\begin{aligned} & \bar{\pi} \\ & \stackrel{1}{\infty} \end{aligned}$ | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
| 2015 | 114.1 | 44.4 | 147.4 | 22.6 | 28.0 | 1363 | 180 | 10.8 | 2,350 | 450 | 20.6 | 10.1 |
| 2016 | 160.2 | 53.5 | 200.3 | 20.0 | 25.1 | 866 | 112 | 23.1 | 3,329 | 519 | 20.8 | 9.7 |
| 2017 | 164.4 | 39.2 | 193.7 | 15.2 | 19.2 | 1600 | 153 | 12.1 | 3,560 | 424 | 21.7 | 10.8 |
| 2018 | 97.8 | 30.4 | 120.6 | 18.9 | 23.7 | 876 | 154 | 13.8 | 1,974 | 336 | 20.2 | 11.0 |
| 2019 | 80.9 | 19.2 | 95.2 | 15.1 | 19.2 | 1121 | 141 | 8.5 | 2,083 | 262 | 25.8 | 13.7 |
| 2020 |  |  | 75.5 | 21.3 | 26.5 | 750 | 118 | 10.1 |  |  | 25.6 | 13.4 |
|  | 59.4 | 21.5 |  |  |  |  |  |  | 1518 | 288 |  |  |
| 2021 | 61.4 | 11.4 | 69.9 | 12.2 | 15.6 | 656 | 87 | 10.7 | 1616 | 149 | 26.3 | 13.1 |


|  |  |  |  |  |  | UWTV abundance estimate |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { 㐫 } \\ & \text { ָ } \end{aligned}$ | millions | millions | millions | \% | \% | millions |  | \% | tonnes | tonnes | gramme | gramme |
| Average 2019-2021 |  |  |  | 16.2 | 20.4 |  |  |  |  |  | 25.9 | 13.4 |



Figure 21.1.1. Nephrops in FU22 (Smalls Grounds). Time-series of the number of Irish vessels reporting landings of Nephrops from FU22 with a >10 threshold.


Figure 21.1.2. Nephrops in FU22 (Smalls Grounds). Combined box and kite plot of vessel power on the Smalls Grounds by year. The blue line indicates the mean.


Figure 21.2.1. Nephrops in FU22 (Smalls Grounds). Landings in tonnes by country.


Figure 21.2.2. Nephrops in FU22 (Smalls Grounds). Fishing effort Kw days for the Irish otter trawl Nephrops directed fleet (30\% of Nephrops weight in total landings).


Figure 21.2.3. Nephrops in FU22 (Smalls Grounds). Sampling levels (numbers) by year and quarter and sample type.


Figure 21.2.4. Nephrops in FU22 (Smalls Grounds). The annual estimated $L_{50}$ with standard error bounds for the on-board retention ogives for samples from the Smalls grounds. Minimum conservation size (MCR) $\mathbf{2 5}$ Carapace Length (CL mm) shown as dashed line.

## Length frequencies for catch (dotted) and landed(solid): <br> Nephrops in FU22



Figure 21.2.5. Nephrops in FU22 (Smalls Grounds). Mean size trends for catches and whole landings by sex over the timeseries.


Figure 21.2.6. Nephrops in FU22 (Smalls Grounds). Sex ratio of the percentage males over the time-series.


Figure 21.2.7. Nephrops in FU22 (Smalls Grounds). Mean weight in catch samples by sex with loess smoother and showing cyclical trends.


Figure 21.2.8. Nephrops in FU22 (Smalls Grounds). Annual mean weights (gr) in the landings and discards.


Figure 21.2.9. Nephrops in FU22 (Smalls Grounds). Contour plots of the krigged density estimates for the UWTV surveys over the time-series.


Figure 21.2.10. Nephrops in FU22 (Smalls Grounds). Time-series of abundance estimates for FU22 (error bars indicate $95 \%$ confidence intervals) and MSY $B_{\text {trigger }}$ is dashed line.


Figure 21.2.11. Nephrops in FU22 (Smalls Grounds). Mean size trends (Carapace length CL mm) for catches by sex from IGFS-WIBTS-Q4 [G7212]. Vertical lines displayed are Minimum Conservation Reference Size $\mathbf{2 5} \mathbf{~ m m}$ Carapace Length (CL) and 35 mm CL.


Figure 21.2.12. Nephrops in FU22 (Smalls Grounds). Mean size trends (Carapace length CL mm) for catches by sex from EVHOE- WIBTS-Q4 [G9527] French survey. Vertical lines displayed are Minimum Conservation Reference Size 25 mm Carapace Length (CL) and 35 mm CL .


Figure 21.3.1. Nephrops in FU22 (Smalls Grounds). Harvest Rate (\% dead removed/UWTV abundance). The dashed and solid lines are the MSY proxy and the harvest rate respectively.

# 22 Norway lobster (Nephrops norvegicus) in divisions 8.a and 8.b, Functional Units 23-24 (northern and central Bay of Biscay) 

The section for Norway lobster (Nephrops norvegicus) in divisions 8.a and 8.b, Functional Units 23-24 (northern and central Bay of Biscay) is found in the 2022 report from the Working Group for the Bay of Biscay and the Iberian Waters Ecoregion (WGBIE).

## 22 Norway lobster (Nephrops norvegicus) in Division 9.a, Functional Unit 30 (Atlantic Iberian waters East and Gulf of Cadiz)

The section for Norway lobster (Nephrops norvegicus) in Division 9.a, Functional Unit 30 (Atlantic Iberian waters East and Gulf of Cadiz) is found in the 2022 report from the Working Group for the Bay of Biscay and the Iberian Waters Ecoregion (WGBIE).

# 23 Plaice (Pleuronectes platessa) in divisions 7.b-c (West of Ireland) 

## Type of assessment in 2020

No assessment was performed.

### 23.1 General

## Stock Identity

Plaice in 7.b are mainly caught by Irish vessels on sandy grounds in coastal areas. Plaice catches in 7.c are negligible. There are two distinct areas in which plaice are caught by Irish vessels in 7.b: an area around Galway Bay and an area in the north of 7.b, which extends into 6.a (the Stags and Broadhaven Ground). During 1995-2000 a large proportion of the 7.bc plaice landings were taken from the Stags Grounds (Rectangles 37D8, 37D9, 37E0 and 37E1). The landings and LPUE in this area have dropped sharply since 2000, in line with a general decrease of LPUE in Division 6.a. Plaice in this area appear to be more linked with $6 . a$ than populations further south. The landings and LPUE on the Aran grounds appear to have been more or less stable since the start of the logbooks' time-series in 1995 (WD 1, WGCSE 2009). It is not known how much exchange there is between plaice on the Aran grounds and those on the Stags ground. The commercial LPUE time-series may not be reflective of overall stock abundance due to changing fishing practices.

## Data

The time-series of official landings is presented in Table 23.1 and Figure 23.1.
Sampling is carried out in Ireland but numbers of samples varies over time due to the low landings levels and varying encounter probability and is not sufficient to generate a time-series of annual length or age distributions. In 2021, 13 size classes were sampled, however, there were only two sampling trips. Figure 23.2 describes the length-frequency distribution of the discard trips, and the contribution of these length classes to hauls and trips.

Table 23.1. Landings of plaice in 7.bc as officially reported to ICES.




Figure 23.1. Landings of plaice in 7.bc as officially reported to ICES (1908-2019).


Figure 23.2. Estimated age distribution of plaice $7 . b c$ in 2021 based on Irish sampling (landings in blue, discards in black).

## 24 Plaice in Division 27.7.a (Irish Sea)

## Type of assessment in 2022

WKIrish3 (ICES, 2017) benchmarked this assessment and choose the SAM model, including estimates of discards-at-age into the catch matrix. A baseline run of the model was performed using discards since 1981 reconstructed according to the medium discard scenario (ICES, 2017).

## ICES advice applicable to 2021

ICES advises that when the MSY approach is applied, catches in 2021 should be no more than 2846 tonnes.
http://ices.dk/sites/pub/Publication\ Reports/Advice/2020/2020/ple.27.7a.pdf

## ICES advice applicable to 2022

ICES advises that when the MSY approach is applied, and assuming that discard rates and fishery selection patterns do not change from the average of the years 2018-2020, total catches in 2022 should be no more than 2747 tonnes.

Advice for 2022 is available at:
https://www.ices.dk/sites/pub/Publication\ Reports/Advice/2021/2021/ple.27.7a.pdf
During the working group the advice for 2022 was updated to 2925 as a result of the revision of the Northern Ireland survey index.

### 24.1 General

## Stock description and management units

The stock assessment area and the management unit are both Division 27.7.a (Irish Sea).

## Management applicable in 2021 and 2022

Management of plaice in Division 27.7.a is by TAC and there is a Minimum Conservation Reference Size (MCRS) of 27 cm in force. The agreed TACs and associated implications for plaice in Division 27.7.a are detailed in the tables below.

## 2021

| Species: | Plaice <br> Pleuronectes platessa | Zone:7a <br> (PLE/07A.) |
| :--- | :--- | :--- |
| Belgium | 62 | Analytical TAC |
| France | 27 | Article 8(2) of this Regulation applies |
| Ireland | 1069 |  |
| The | 19 |  |
| Netherlands | 1177 |  |
| Union | 2855 |  |
| United |  |  |
| Kingdom |  |  |
| TAC |  |  |

(Source: Council Regulation (EU) 2021/1239, ANNEX IA)

## 2022

| Species: | Plaice <br> Pleuronectes platessa | Zone:7a <br> (PLE/07A.) |
| :--- | :--- | :--- |
| Belgium | 60 | Analytical TAC |
| France | 26 | Article 8(2) of this Regulation applies |
| Ireland | 1031 |  |
| Netherlands | 19 |  |
| Union | 1136 |  |
| United Kingdom | 1404 |  |
| TAC | 2747 |  |

(Source: Council Regulation (EU) 2022/515, ANNEX IA)

## The fishery in 2021

National landings data reported to ICES and Working Group estimates of total landings are given in Table 24.1. A summary by gear is given below.

| Catch (2021) | Landings |  | Discards |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |

The TAC for 2021 was 2846 tonnes and the working group estimate of landings in 2021 was 276 tonnes. The poor uptake of the quota is not a consequence of an inability to catch sufficient quantities of plaice greater than the MCRS but rather is most likely due to the limited market demand and poor value of the catch.

Landings (based on working group estimates) by the Irish, NI, UK and Belgian fleets comprised approximately $38 \%, 3.6 \%, 22 \%$ and $36 \%$ respectively of total landings in 2021. The landings of plaice are mainly split between beam trawlers ( $68 \%$; primarily Belgian vessels then Irish vessels) targeting sole, and otter trawlers ( $31 \%$; Irish and UK vessels). Historically, otter trawling was dominated by UK vessels fishing for whitefish, but in recent years, many vessels have switched to target Nephrops (Figure 24.1). Otter trawlers from Ireland and N. Ireland typically target Nephrops in the western Irish Sea.
High levels of discarding are known to occur in all fisheries that catch plaice in the Irish Sea (see Figures 24.4 and 24.5).
A general description of the fishery can be found in the stock annex and also in 'Other Relevant Data' section below.

### 24.2 Data

## Landings

National landings data reported to ICES and Working Group estimates of total landings are given in Table 24.1. The working group procedures used to determine the total international landings numbers- and weights-at-age are documented in the stock annex. As a result of increased rates of discarding, landed numbers-at-age for the younger ages (ages 2 to 4 ) have declined more rapidly over the last two decades than landings of older fish (Figure 24.2a). Age compositions of landed fish are available for Belgian beam trawl, Irish beam and otter trawl and English otter trawl (Figure 24.5).

## Discards

Discard sampling has been conducted by the UK(E\&W) since 2002 and by Ireland since 1993; Northern Ireland has collected data from 1996 (but not between 2003 and 2005), and Belgium since 2003. Age compositions of discarded fish are available for Belgian beam trawl and Irish otter trawl, discard estimates are also available for Northern Irish otter trawl. Discard rates for unsampled fleets are taken from the sampled fleets separately for fisheries targeting demersal fish and Nephrops. For 2021 however, discard estimates for the English fleets are calculated using the English Nephrops average discard rate (2017-2019) due to concerns that the discarding practices of Northern Irish and Irish Nephrops fleets were not applicable to this fleet, and would have implied an unrealistically large amount of discards (Figure 24.4). Age compositions of discarded fish are available for Belgian beam trawl (used for gears targeting demersal fish) and Irish otter trawl (used for gears targeting Nephrops) (Figure 24.5).
WKFLAT (ICES, 2011) first estimated total international discards-at-age and introduced them to the assessment of the stock for the first time. Due to limitations in the data available by gear type, discards for Ireland, France and Northern Ireland, for the years 2004-2011 were raised using UK estimates on the basis of equivalent gear types. A raising factor based on tonnages landed for these countries was calculated and applied to the UK(E\&W) estimates of discard numbers. Finally, these estimates were added to those calculated for Belgium to give estimates of total international discard numbers-at-age.

There is a considerable historic time period (1981-2003) for which no international raised discard estimates are available. The method for reconstructing discards prior to 2004 is based on sizevarying discard rates and is documented in Annex 4 of the WKIrish3 report (ICES, 2017).

Since 2012, catch data (landings and discards) are available from InterCatch disaggregated by country and fleet. Total international discards are raised from available discards data.

The total discard estimates (Table 24.1, Figure $24.2 b$ ) confirm the significant proportion of discarding that occurs in the fishery, which has increased in time. Since 2004, the majority of the catch has been discarded $(60 \%$ and $81 \%$ average discard in weight and in numbers respectively, since 2004).

## Biological

Landings numbers-at-age are given in Table 24.5 and plotted in Figure 24.2a. Weights-at-age in the landings are given in Table 24.6. Discard weights-at-age are given in Table 24.7 and weights-at-age in the stock in Table 24.8. The history of the derivation of the landings weights and stock weights used in this assessment is described in the stock annex.

Mean weight-at-age in the landings and survey data indicate declines in both sexes throughout the Irish Sea since 1993 so that plaice at ages $\leq 4$ are typically below MCRS.

## Surveys

All available tuning data are shown in Tables 24.2, 24.3 ( $a$ and $b$ ) and 24.4. Due to inconsistencies in the available commercial tuning fleets, Irish Sea plaice assessments since 2004 have only included the UK (E\&W) beam trawl survey (UK (E\&W)-BTS-Q3) (B6596) and the two NIGFS-WIBTS spawning biomass indices based on ground fish surveys (NIGFS-WIBTS-Q1 (G7144) and NIGFS-WIBTS-Q4 (G7655)). For more information see WGNSDS (ICES, 2004). The UK (E\&W)-BTS-Q3 index was revised by WKFLAT 2011 to include stations in the western Irish Sea and in St George's Channel. A second revision was conducted in 2017 to correct for some inconsistency in the index calculation This revision did not substantially change the trend of the biomass index (see WD Cambiè and Earl, 2017 in WGCSE 2017 report).

Reviews of the UK (E\&W)-BTS-Q3 mean standardised CPUE trends have indicated that the survey has good internal consistency in monitoring trends across the stock area. For the entire Irish Sea, the biomass index calculated from the UK (E\&W)-BTS-Q3 (Figure 24.3, right) indicates two periods of upwards trend, 1993-2003 and from 2007-2015. It is however, detected to have dropped from 2016. An increase of numbers in older ages is observed until around 2015, followed by a steep decline (Figure 24.3, left). The NIGFS-WIBTS surveys show similar increases in biomass between 1993 and 2003 and then a further increase subsequently until most recent years.

In 2022, an error was discovered in the calculation of the tuning index from the NIGFS-WIBTS (Autumn and Spring), going back four years for the Autumn series and three years for the Spring series. The most recent datapoints were not calculated in the same units as previously, and so had to be corrected. The revision increased these data points by a factor of 3 . An update of the 2021 assessment using the corrected data led to the SSB at the start of the interim year being revised up by $13 \%$ and a minor improvement to the model fit.
The NIGFS-WIBTS survey strata can be disaggregated into western (Strata 1-3) and eastern (Strata 4-7) subareas, where the subareas are divided by the deep trench that runs roughly northsouth to the west of the Isle of Man (Figure 24.6, Tables 24.3a and b).

The SSB of plaice in the Irish Sea was also independently estimated using the Annual Egg Production Method (AEPM), according to Armstrong et al., 2001 methodology.

| Year | SSB (tonnes) | Catch/SSB harvest rate |  |
| :--- | :---: | :---: | :---: |
| 1995 | 9081 |  |  |
| 2000 | 13303 | 15.16 |  |
| 2006 | 14417 | 14352 | 19.5 |
| 2008 | 15071 |  |  |

Catch (discards available from 2004) to egg survey biomass ratios indicate historically that the plaice in the Irish Sea has been lightly exploited. Splitting the SSB estimates from the AEPM into eastern and western Irish Sea areas (Figure 24.7) also indicates that the perceived increase in plaice biomass is due to increased production in the eastern Irish Sea only (for more details see stock annex).

In summary, the UK (E\&W)-BTS-Q3 in September, the NIGFS-WIBTS-Q4 index in October (but not NIGFS-WIBTS-Q1 March), and the AEPM indicate a sustained increasing trend in biomass in the eastern Irish Sea, but this rise does not appear to extend across the deep channel to plaice in the western Irish Sea (Figure 24.7).

## Commercial CPUE

Age-based tuning data available for this assessment, comprise three commercial fleets: the UK(E\&W) otter trawl fleet (UK(E\&W) OTB, from 2008), the UK(E\&W) beam trawl fleet (UK(E\&W) BT, from 1989) and the Irish otter trawl fleet (IR-OTB, from 1995). Due to inconsistencies in the available tuning fleets, Irish Sea plaice assessments since 2004 have omitted these indices. For more information, see WGNSDS 2004. The effort and catch by these commercial fleets has been very low in recent years and the CPUE data are no longer considered informative.

## Other relevant data

Table 24.2 and Figure 24.1 show that effort levels have decreased since 2002 for the majority of fleets. Both the UK otter and beam trawl fleets are close to their lowest recorded effort levels in time-series extending back to 1972 and 1983 respectively. Effort by UK Nephrops trawlers has greatly increased in the years 2006-2014 but has decreased in the last years. However, this fleet is now the dominant UK fleet in terms of hours fished in 27.7.a. Belgian vessels operating in Division 7 typically move in and out of the Irish Sea, depending on the season, from specifically the Bristol Channel and Celtic Sea, the Bay of Biscay and the southern North Sea.

Since 2013, a problem with the gear effort information (000s hours fished) reported for the UK (E\&W) commercial beam trawl fleet has been registered. Effort information from this fleet is largely missing as a result of a larger component of the fleet using the EU electronic logbook system to report its activities. Gear effort information reporting has not been mandatory with this system to date. As a result, few trips reported their gear effort information rendering the overall effort reported and resulting LPUE unusable. However, an initial inspection of an alternate effort indicator for this gear (days fished) suggests that UK beam trawl effort in 2013, 2014, 2015, 2016, and 2017 is at the level observed in 2012. The otter trawl fleet effort reporting was unaffected by this as these vessels were not reporting their landings via this method in these years.

### 24.3 Historical stock development

Model: Age-based analytical assessment (State-space Assessment Model, SAM) that uses landings and discards (Nielsen and Berg, 2014).

Software: R version 4.0.2 with additional packages (version in parenthesis):
stockassessment (0.11.0); FLCore (2.6.18); reshape (0.8.9); ggplot2 (3.3.5); Cairo (1.5.15); doParallel (1.0.17); TMB (1.8.1); devtools (2.4.3).

## Model options chosen

The AP model (Aarts and Poos, 2009) was replaced by SAM. WGCSE (ICES, 2016) agreed that the AP model was not the definitive assessment tool for Irish Sea plaice but a temporary solution to the fitting of datasets which included recent discards estimates but for which historic discard information was not available. Reconstructed values of historic discards (prior 2004) were provided in the WKIrish3 (ICES, 2017). The SAM model incorporates the estimated historic discards and is used to run the assessment since 2017.

The model runs were performed using the R package 'stockassessment'. Settings for this update stock assessment are given in the table below. The update assessment follows the same procedure as in the stock annex (ICES, 2017). A baseline run of the model was performed using discards since 1981 reconstructed according to the medium discard scenario (ICES, 2017). Discard survival was set at $40 \%$, and natural mortality followed a Lorenzen curve, scaled to 0.12 .

## Input data types and characteristics

Commercial catch-at-age data. Discards values available from 2004. Estimates of discards reconstructed for 1981-2003 (ICES, 2017). Only the dead fraction of discards (0.6) is accounted for in the model. Three survey indices (UK (E\&W)-BTS-Q3, NIGFS-WIBTS-Q1, and NIGFS-WIBTSQ4); fixed maturity ogive; natural mortality constant over years and different across ages.

## Final update assessment

WKIrish3 (ICES, 2017) benchmarked this assessment and included estimates of discards-at-age into the catch matrix.

The assessment settings are shown in the following table. Historic settings are given in the stock annex.

| Assessment year |  | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Assessment model |  | SAM | SAM | SAM | SAM | SAM |
| Tuning fleets | UK <br> (E\&W)- <br> BTS-Q3 | Survey omitted | Survey omitted | Survey omitted | Survey omitted | Survey omitted |
|  | Extended <br> UK <br> (E\&W)- <br> BTS-Q3 | $\begin{aligned} & \text { 1993-2017, ages } \\ & 1-7 \end{aligned}$ | $\begin{aligned} & \text { 1993-2018, ages } \\ & 1-7 \end{aligned}$ | $\begin{aligned} & \text { 1993-2019, ages } \\ & 1-7 \end{aligned}$ | $\begin{aligned} & \text { 1993-2019, ages } \\ & 1-7 \end{aligned}$ | $\begin{aligned} & \text { 1993-2021, } \\ & \text { ages 1-7 } \end{aligned}$ |
|  | UK(E\&W) BTS Mar | Survey omitted | Survey omitted | Survey omitted | Survey omitted | Survey omitted |
|  | UK(E\&W) OTB | Series omitted | Series omitted | Series omitted | Series omitted | Survey omitted |
|  | $\begin{aligned} & \text { UK(E\&W) } \\ & \text { BT } \end{aligned}$ | Series omitted | Series omitted | Series omitted | Series omitted | Survey omitted |
|  | IR-OTB | Series omitted | Series omitted | Series omitted | Series omitted | Survey omitted |
|  | NIGFS- <br> WIBTS- <br> Q1 | 1992-2017 | 1992-2018 | 1992-2019 | 1992-2020 | 1992-2021 |
|  | NIGFS- <br> WIBTS- <br> Q4 | 1992-2017 | 1992-2018 | 1992-2019 | 1992-2020 | 1992-2021 |
| Selectivity model |  | Correlated random walk | Correlated random walk | Correlated random walk | Correlated random walk | Correlated random walk |
| Discard fraction |  | Estimated by WKIrish3 | Estimated by WKIrish3 | Estimated by WKIrish3 | Estimated by WKIrish3 | Estimated by WKIrish3 |
| Landings N at age |  | $\begin{aligned} & \text { 1981-2017, ages } \\ & 1-8+ \end{aligned}$ | $\begin{aligned} & \text { 1981-2018, ages } \\ & 1-8+ \end{aligned}$ | $\begin{aligned} & \text { 1981-2019, ages } \\ & 1-8+ \end{aligned}$ | $\begin{aligned} & \text { 1981-2020, ages } \\ & 1-8+ \end{aligned}$ | $\begin{aligned} & \text { 1981-2021, } \\ & \text { ages 1-8+ } \end{aligned}$ |
| Discards N at age |  | $\begin{aligned} & \text { 1981-2017, ages } \\ & 1-8+ \end{aligned}$ | $\begin{aligned} & \text { 1981-2018, ages } \\ & 1-8+ \end{aligned}$ | $\begin{aligned} & \text { 1981-2019, ages } \\ & 1-8+ \end{aligned}$ | $\begin{aligned} & \text { 1981-2020, ages } \\ & 1-8+ \end{aligned}$ | $\begin{aligned} & \text { 1981-2021, } \\ & \text { ages 1-8+ } \end{aligned}$ |

The estimated selectivity patterns split into the landed and discarded components are shown in Figure 24.8. Until early 1990s, the landings selectivity had the highest values for fish aged 4 (indicating that 4-year aged fish were selected). This selectivity shifted to age 5 in late the 1990s and early 2000s, due to the increase of the MCRS in 1998 (from 250 mm to 270 mm ). Since late 2000s landings gradually fell over time to very low values relative to the discard pattern, which became dominant and expanded to the older aged fish during the most recent years.

The catchability of the UK (E\&W)-BTS-Q3 survey is elevated for ages 1 and 2 and reflects the nature of the survey, which was designed as a recruit index (Figure 24.9).

Diagnostic output from the SAM model is shown in Figure 24.10. In the catch residuals, negative values are apparent in ages $8+$ from 1998. A year effect in 2004 is present in the UK(E\&W)-BTSQ3 residuals (which is the first year for which discard data are available). A pattern of negative
residuals between 2004 and 2009 is present in the residuals of the NIGFS-WIBTS due to large fluctuations in the SSB indices, which are due potentially to variable catchability of the survey.

Recruitment is estimated to be fluctuating without an overall trend until 2015, and then estimated at its lowest values in 2017-2021. The standardised values of the recruitment estimated by the SAM model and the standardised value of age 1 from the UK-BTS survey are characterised by similar pattern, demonstrating consistency in the model estimates (Figure 24.11).

The estimated SSB from the SAM model shows an increasing trend from 1995 until 2004-2005, followed by a drop in 2006 and 2007. This change in SSB trend from 2004 is probably due to the inclusion of more reliable discards values since 2004, when international raised discard estimates became available. Since 2012, SSB has increased reaching the highest value of the whole timeseries in 2016, followed by a significant in estimated SSB since then. The SSB trends are largely in agreement with independent SSB estimates from the Annual Egg Production Method (AEPM), up to the most recent estimate in 2010, as well as showing a similar trend to the survey data used in the assessment (NIGFS-WIBTS-Q1 and -Q4; UK(E\&W)-BTS-Q3, Figure 24.12).

Estimates of numbers-at-age in the landings, discards and population, and fishing mortality numbers-at-age are given in Tables 24.9-24.12. A summary plot for the SAM assessment is shown in Figure 24.13 and the time-series estimates for $\mathrm{Fbar}^{2}$, SSB and recruitment are given in Table 24.13.

## Comparison with previous assessments

In 2017, the Aarts and Poos model was replaced by the state-space assessment model (SAM). The assessment used the Lorenzen M scaled to 0.12 , and the most recent maturity ogive for the survey.

The methodology provided is as robust as possible and does not currently appear to suffer from a serious retrospective pattern (Figures 24.14 and 24.15). The ten assessment model configurations compared in WKIrish3 (ICES, 2017) perform similarly in terms of temporal trends in SSB, recruitment, catch and Fbar . Small retrospective bias in SSB in 2004 likely resulted from the introduction of discards estimates based on samples collected from that year (prior to 2004, discards estimates are reconstructed values based on size-varying discard rates). A Mohn's rho analysis for a five-year peel resulted in values of $2.8 \%$ for recruitment, $2.2 \%$ for SSB and $-3.8 \%$ for $\mathrm{F}_{\mathrm{bar}}$.

## State of the stock

Trends in $\mathrm{Fbar}_{\mathrm{bar}}$ SSB, recruitment and catch, for the full time-series, are shown in Table 24.13 and Figure 24.13. The assessment consistently estimates that fishing mortality declined from high levels in the 1980s and early 1990s to very low levels, having been $<0.1$ since 2013 . Since 2012, SSB has increased reaching the highest value of the whole time-series in 2016, whereas it has slightly decreased in 2017. Estimated recruitments are highly variable. An increasing trend was present until 2015 although it seems to have dropped to the lowest values in 2017-2020. Catch has decreased to low levels and, since 2006, the majority of the catch has been discarded ( $60 \%$ in weight and $81 \%$ and number respectively, averaged since 2004).

### 24.4 Short-term projections

Forecasting takes the form of short-term stochastic projections. A total of 1000 samples are generated from the estimated distribution of survivors. These replicates are then simulated forward according to model and forecast assumptions (see table below), using the usual exponential decay equations, but also incorporating the stochastic survival process (using the estimated survival standard deviation) and subject to different catch-options scenarios. Recruitment in the
intermediate year (2022) was taken as the median from a distribution about the assessment estimate. Estimates of recruitment for intermediate year and subsequent years were resampled from the 2015-2021 year classes, reflecting recent low levels of recruitment. These re-sampled recruitments are used in SAM forecasts in order to evaluate future stock dynamics.

| Initial stock size | Starting populations are simulated from the estimated distribution at the start of the inter- <br> mediate year (including covariances) |
| :--- | :--- |
| Maturity | Average of final three years of assessment data of final three years of assessment data |
| Natural mortality | Both taken as zero |
| F and M before spawn- <br> ing | Assumed to be the same as weight-at-age in the catch final three years of assessment data |
| Weight at age in the <br> catch | Fishing mortalities taken as a three-year average |
| Weight at age in the <br> stock | Recruitment for the intermediate year onwards is sampled, from 2015 to the final year of <br> catch data |
| Stock recruitment <br> model used | An average of final three years of landing fractions are used in the forecast period <br> Procedures used for <br> splitting projected <br> catches |

F estimates 2017-2019 has fluctuated around similar values, with further decline in 2020. F status quo, Fsq, has been estimated by averaging the F over 2019-2021 (0.064).

A full management options table is provided in Table 24.15, based on the intermediate year assumption in Table 24.14. Note that the values that appear in the catch scenarios are medians from the distributions that result from the stochastic forecast. Implementing the management plan for this stock with $\mathrm{F}_{\mathrm{MSY}}=0.196$ leads to a total catch of 2039 t ( 967 t of landings and 1072 t of discards including dead and survivors) in 2023 and SSB of 12629 t in 2024.

### 24.5 Medium-term projections

There are no medium-term projections for this stock.

### 24.6 MSY explorations

The reference points for this stock were estimated in 2018 (ICES, 2018) as ICES request for EU western waters stocks and are presented in the table below. in 2021, ICES changed the basis for $\mathrm{F}_{\mathrm{pa}}$ to $\mathrm{F}_{\mathrm{p} .05 \text {, and }}$ the updated $\mathrm{F}_{\mathrm{pa}}$ value is shown in the table below.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | $8757$ <br> tonnes | Lower 5th percentile of B $\mathrm{F}_{\text {MSY }}$ | ICES (2018) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.196 | Stochastic simulations with segmented regression from the entire time-series (1981-2017) | ICES (2018) |
|  | $\mathrm{F}_{\text {MSY lower }}$ | 0.133 | F at 95\% MSY (below FMSY), based on simulation using a segmented regression stock-recruitment relationship (EqSim) | ICES (2018) |
|  | $\mathrm{F}_{\text {MSY upper }}$ | 0.293 | F at 95\% MSY (above FMSY), based on simulation using a segmented regression stock-recruitment relationship (EqSim) | ICES (2018) |
| Precautionary approach | $\mathrm{Blim}_{\text {lim }}$ | $3958$ <br> tonnes | $\mathrm{B}_{\text {loss }}=$ minimum SSB observed | ICES (2018)) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | $5294$ <br> tonnes | $\mathrm{B}_{\lim } \times \exp (1.645 \times \sigma) ; \sigma=0.177$ | ICES (2018) |
|  | $\mathrm{F}_{\text {lim }}$ | 0. 50 | F with $50 \%$ probability of $\mathrm{SSB}<\mathrm{Bl}_{\text {lim }}$ | ICES (2018) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.403 | $\mathrm{F}_{\mathrm{p} .05}$; the F that leads to $\mathrm{SSB} \geq \mathrm{B}_{\mathrm{lim}}$ with $95 \%$ probability | ICES (2018) |
| Management plan | SSB ${ }_{\text {mgt }}$ | Not applicable |  |  |
|  | $F_{\text {mgt }}$ | Not applicable |  |  |

## Yield per Recruit analysis

There are no yield per recruit analyses for this stock.

### 24.7 Management plans

There are no management plans for this stock.

### 24.8 Uncertainties and bias in assessment and forecast

The assessment was benchmarked in 2017 (WKIrish3), which resulted in the SAM model being fitted using catches based on reconstructed estimates of discards prior to 2004. This discard reconstruction introduces additional uncertainty in the model. The model estimates of stock development since 2004 are more reliable as based on direct discard estimates. The SAM model considered only the dead portion of the discards ( $60 \%$ ), but in the forecast the estimates are raised to include the surviving discards. The Mohn's rho measure of retrospective bias for this assessment is low (Section 24.3).

The assessment indicates that recruitment and $F$ have both been falling in recent years, and as a result the average age of catches has been increasing. An increasing amount of the stock is contained within the modelled plusgroup ( $47 \%$ in the last five years is age $8+$ ). Consequently, the
assessment and forecast have increased uncertainty and a pattern of retrospective adjustment of terminal year SSB downwards is seen in the recent history of the assessment.

### 24.9 Recommendations for next benchmark

There is evidence of substantial substock structure and incorporating information about the differences in growth and maturity between the east and west sides of the Irish Sea, as well as by sex should be explored.

Incorporating data on changes in maturity and natural mortality over time, linked to the decreasing in weights-at-age observed in survey data, should also be considered. There is evidence of a decline in weight-at-age from the commercial landings data and survey data. The UK(E\&W)-BTS-Q3 survey data also indicate declines in length-at-age and maturity-at-age.

Creating age-based indices for the NI groundfish surveys would improve the assessment.
Ecosystem information ought to be explored.

| Type | Problem/Aim | Work required | Data required | Expertise required |
| :---: | :---: | :---: | :---: | :---: |
| Sampling | The split between OTB and BTT has changed, and sample raising may not adequately reflect the changed split | Review consistency of sample raising to ensure the change of OTB/BTT is accurately and consistently reflected in the raised samples | Data already available in InterCatch | Catch sampling expertise |
| Assessment method | The assessment indicates that recruitment and $F$ have both been falling in recent years, and as a result the average age of catches has been increasing. An increasing amount of the stock is contained within the modelled plusgroup (47\% in the last five years is age $8+$ ). Consequently, the assessment and forecast have increased uncertainty and a pattern of retrospective adjustment of terminal year SSB downwards is seen in the recent history of the assessment. | Recompile age distributions with a higher plusgroup, test effect of different Catchability assumptions in this age group. Consider whether $F_{\text {bar }}$ age range needs changing. Possible recalculation of reference points | Landings data by age, as disaggregated as possible. Should be available post 2004 in InterCatch, but historic data availability unknown | Historic catch age composition raising |
| Other issues | Fits to NIGFS indices use SSB indices, assuming constant selectivity for all age/length | Explore whether age/length compositions can inform the selectivity of the survey and whether this can be included in the assessment | Survey age/length compositions | Survey index compilation experts |


| Type | Problem/Aim | Work required | Data required | Expertise required |
| :---: | :---: | :---: | :---: | :---: |
| Biological parameters | Natural mortality and maturity may be connected with size which has varied substantially over time and between parts of the stock. | Investigate whether time varying biological parameters can be derived and used in this stock | stock size data, relationships between M and stock size, relationships between maturity and stock size. Has the catch split changed between East and West of the area, and does this affect average M and maturity? |  |

### 24.10 Management considerations

The high level of discarding in this fishery indicates a mismatch between the minimum landing size and the mesh size of the gear being used. Any measures that effect a reduction in discards will result in increased future yield. However, the market demand for plaice is poor and small plaice are particularly undesirable. Strong year effects are seen in the discard data and these are likely due to spatial structure in the stock. Spatial management of fleets in the Irish Sea may reduce the discarding of plaice.

The overall state of the stock is consistently estimated to have low fishing mortality and high spawning biomass. Therefore, the stock is considered to be within safe biological limits.

Discarding has increased throughout the period in which data are available, while landings of plaice have decreased, even though the TAC is not restrictive. Effort has decreased in fisheries targeting plaice (including UK(E\&W) and Belgian beam-trawl fisheries, and UK(E\&W) and Irish otter trawl fisheries targeting demersal fish). In contrast, effort by the UK(E\&W) Nephrops fleet has increased, however, this is still small in comparison to effort by the Irish Nephrops fleet. The main Nephrops grounds are located in the western Irish Sea, where relatively small plaice are found. Technical measures to mitigate discarding by all Nephrops fleets could include the use of sorting grids: gear selectivity trials and monitoring from four Irish Nephrops trawlers using grids since 2009 indicate a potential reduction in fish discarding by $75 \%$ (BIM, 2009).

### 24.11 References

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Table 24.1. Plaice in Division 7.a. History of official landings and ICES estimates of discards. Weights are in tonnes.

| Year | $\begin{aligned} & \frac{\varepsilon}{工} \\ & \frac{10}{00} \\ & \frac{0}{0} \end{aligned}$ | $\begin{aligned} & \text { 쁜 } \\ & \text { 든 } \end{aligned}$ | $\begin{aligned} & \mathbf{D} \\ & \underline{\widetilde{T}} \\ & \underline{\underline{N}} \end{aligned}$ |  |  |  | $\begin{aligned} & \overline{0} \\ & \frac{\pi}{5} \\ & \stackrel{0}{0} \\ & \dot{y} \\ & \check{j} \end{aligned}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 332 | 13 | 547 | - | 1082 | 14 | 63 | 2051 |  |
| 1995 | 327 | 10 | 557 | - | 1050 | 20 | 60 | 2024 |  |
| 1996 | 344 | 11 | 538 | 69 | 878 | 16 | 18 | 1874 |  |
| 1997 | 459 | 8 | 543 | 110 | 798 | 11 | 25 | 1954 |  |
| 1998 | 327 | 8 | 730 | 27 | 679 | 14 | 18 | 1803 |  |
| 1999 | 275 | 5 | 541 | 30 | 687 | 5 | 23 | 1566 |  |
| 2000 | 325 | 14 | 420 | 47 | 610 | 6 | 21 | 1443 |  |
| 2001 | 482 | 9 | 378 | - | 607 | 1 | 11 | 1488 |  |
| 2002 | 636 | 8 | 370 | - | 569 | 1 | 7 | 1591 |  |
| 2003 | 628 | 7 | 490 | - | 409 | 1 | 9 | 1544 |  |
| 2004 | 431 | 2 | 328 | - | 369 | 0 | 4 | 1134 | 1031 |
| 2005 | 566 | 9 | 272 | - | 422 | 0 | 1 | 1270 | 1210 |
| 2006 | 343 | 2 | 179 | 0 | 413 | 0 | 0 | 937 | 1254 |
| 2007 | 194 | 2 | 194 | 0 | 412 | 0 | - | 802 | 1744 |
| 2008 | 157 | 2 | 102 | 0 | 300 | 1 | 1 | 563 | 1268 |
| 2009 | 197 | 0 | 73 | 0 | 184 | 1 | 2 | 457 | 1132 |
| 2010 | 138 | 0 | 89 | 0 | 147 | 0 | 3 | 377 | 2561 |
| 2011 | 332 | 0 | 118 | 0 | 146 | 0 | 0 | 596 | 603 |
| 2012 | 236 | 0 | 107 | 0 | 164 | 0 | 0 | 507 | 1010 |
| 2013 | 144 | 0 | 103 | 0 | 92 | 0 | 0 | 339 | 725 |
| 2014 | 100 | 0 | 123 | 0 | 59 | 0 | 0 | 282 | 943 |
| 2015 | 115 | 0 | 244 | 0 | 80 | 0 | 0 | 439 | 572 |
| 2016 | 82 | 0 | 541 | - | 56 | - | - | 679 | 437 |
| 2017 | 77 | 0 | 446 | - | 62 | 1 | - | 585 | 852 |
| 2018 | 53 | 0 | 316 | - | 66 | - | - | 435 | 395 |
| 2019 | 168 | 0 | c | - | 57 | 0 | - | $255^{\text {c }}$ | 537 |
| 2020* | 84 | - | 177 | - | 70 | 2 | - | 333 | 271 |
| 2021* | 103 | 0 | 107 | - | 70 | 1 | - | 281 | 392 |

[^14]Table 24.2. Irish Sea plaice: English standardised LPUE and effort, Belgian beam trawl LPUE and effort and Irish otter trawl LPUE and effort series.

| Year | CPUE |  |  | LPUE |  |  |  |  |  |  | Effort |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | UK(E\&W) Beam trawl survey ${ }^{1}$ |  |  | Otter ${ }^{3}$ <br> Trawl | UK (E\&W) ${ }^{2}$ |  | Beam ${ }^{4}$ <br> Trawl | $\begin{aligned} & {\frac{\text { Belgian }^{5}}{}}_{\text {Beam }}^{\text {Trawl }} \\ & \end{aligned}$ | Irish ${ }^{7}$ |  | UK (E\&W) |  |  |  |  | $\begin{aligned} & \frac{\text { Belgian }}{\text { Beam }} \\ & \text { Trawl } \end{aligned}$ | Irish ${ }^{9}$ |  |
|  | March | September | September |  | Otter ${ }^{4}$ | Beam ${ }^{3}$ |  |  | Otter | Beam | Otter ${ }^{3}$ | Otter ${ }^{4}$ | Beam | Beam ${ }^{4}$ | Nephrops ${ }^{3}$ |  | Otter | Beam |
|  |  | Prime only | Extended |  | Trawl | Trawl |  |  | Trawl | Trawl | Trawl | Trawl | Traw T | Trawl | Trawl |  | Trawl | Trawl |
| 1972 |  |  |  | 6.96 |  |  |  | 9.8 |  |  | 128.4 |  |  |  |  | 6.8 |  |  |
| 1973 |  |  |  | 6.33 |  |  |  | 9.0 |  |  | 147.6 |  |  |  |  | 16.5 |  |  |
| 1974 |  |  |  | 7.45 |  |  |  | 10.4 |  |  | 115.2 |  |  |  |  | 14.2 |  |  |
| 1975 |  |  |  | 7.71 |  |  |  | 10.7 |  |  | 130.7 |  |  |  |  | 16.2 |  |  |
| 1976 |  |  |  | 5.03 |  |  |  | 5.8 |  |  | 122.3 |  |  |  |  | 15.1 |  |  |
| 1977 |  |  |  | 4.82 |  |  |  | 5.3 |  |  | 101.9 |  |  |  |  | 13.4 |  |  |
| 1978 |  |  |  | 6.77 |  | 4.88 |  | 6.9 |  |  | 89.1 |  | 0.9 |  |  | 12.0 |  |  |
| 1979 |  |  |  | 7.18 |  | 15.23 |  | 8.0 |  |  | 89.9 |  | 1.7 |  |  | 13.7 |  |  |
| 1980 |  |  |  | 8.24 |  | 8.98 |  | 8.6 |  |  | 107.0 |  | 4.3 |  |  | 20.8 |  |  |
| 1981 |  |  |  | 6.87 |  | 4.91 |  | 7.1 |  |  | 107.1 |  | 6.4 |  |  | 26.7 |  |  |
| 1982 |  |  |  | 4.92 |  | 1.77 |  | 4.4 |  |  | 127.2 |  | 5.5 |  |  | 21.3 |  |  |
| 1983 |  |  |  | 5.32 | 1021 | 3.08 | 0 | 7.8 |  |  | 88.1 | 1716.5 | 2.8 | 0 |  | 18.5 |  |  |
| 1984 |  |  |  | 7.77 | 1472 | 6.98 | 810 | 6.8 |  |  | 103.1 | 7932.1 | 4.1 | 263 |  | 13.6 |  |  |
| 1985 |  |  |  | 9.97 | 1946 | 25.70 | 5487 | 8.8 |  |  | 102.9 | 6930.8 | 7.4 | 428.1 |  | 21.9 |  |  |
| 1986 |  |  |  | 9.27 | 1597 | 4.21 | 753 | 8.7 |  |  | 90.3 | 6693.2 | 17.0 | 1122.9 |  | 38.3 |  |  |
| 1987 |  |  |  | 7.20 | 1479 | 3.57 | 963 | 8.2 |  |  | 130.6 | 9008.9 | 22.0 | 1178.5 |  | 43.2 |  |  |
| 1988 |  | 392 |  | 5.02 | 1060 | 3.05 | 743 | 6.3 |  |  | 132.0 | 8292.4 | 18.6 | 1019.2 |  | 32.7 |  |  |
| 1989 |  | 253 |  | 5.51 | 1109 | 13.59 | 2559 | 6.2 |  |  | 139.5 | 16161.4 | 25.3 | 1344.5 |  | 36.7 |  |  |
| 1990 |  | 239 |  | 5.93 | 1074 | 12.02 | 3011 | 7.2 |  |  | 117.1 | 7724.5 | 31.0 | 1473.1 |  | 38.3 |  |  |
| 1991 |  | 157 |  | 4.79 | 916 | 10.56 | 2807 | 7.5 |  |  | 107.3 | 7081.1 | 25.8 | 1211.3 |  | 15.4 |  |  |
| 1992 |  | 188 |  | 4.20 | 719 | 9.99 | 2303 | 11.9 |  |  | 96.8 | 6671.8 | 23.4 | 908.1 |  | 23.0 |  |  |
| 1993 | 91 | 235 | 149 | 3.97 | 667 | 9.50 | 2220 | 5.0 |  |  | 78.9 | 6013.1 | 21.5 | 826.9 |  | 24.4 |  |  |
| 1994 | 128 | 225 | 132 | 4.90 | 770 | 7.79 | 1020 | 9.2 |  |  | 43.0 | 3060 | 20.1 | 1451.6 | 0 | 31.6 |  |  |
| 1995 | 134 | 169 | 109 | 5.08 | 806 | 7.69 | 1001 | 9.5 | 3.2 | 17.3 | 43.1 | 3357 | 20.9 | 1429.4 | 0 | 27.1 | 80.1 | 8.5 |
| 1996 | - ${ }^{6}$ | 210 | 111 | 5.37 | 732 | 12.96 | 2587 | 11.8 | 4.1 | 19.0 | 42.2 | 3085.1 | 13.3 | 894.3 | 0 | 22.2 | 64.7 | 6.2 |
| 1997 | 147 | 262 | 148 | 5.25 | 662 | 7.66 | 944 | 13.9 | 3.1 | 13.7 | 39.9 | 2903.3 | 10.8 | 784.4 | 0 | 29.3 | 92.0 | 9.9 |
| 1998 | 113 | 249 | 146 | 5.00 | 657 | 5.66 | 766 | 12.3 | 3.7 | 22.3 | 36.9 | 2620.6 | 10.4 | 696 | 0 | 23.8 | 93.5 | 11.5 |
| 1999 | ${ }^{6}$ | 264 | 151 | 5.38 | 632 | 7.76 | 895 | 7.1 | 2.3 | 23.2 | 22.9 | 1803.5 | 11.0 | 778.9 | 0 | 37.2 | 109.7 | 14.7 |
| 2000 | $-6$ | 357 | 169 | 5.02 | 828 | 13.04 | 1773 | 7.8 | 2.0 | 13.8 | 27.0 | 2034.9 | 6.3 | 410.7 | 0 | 27.0 | 82.6 | 11.4 |
| 2001 |  | 281 | 147 | 3.35 | 539 | 8.33 | 1017 | 9.2 | 2.9 | 14.0 | 33.0 | 2352.9 | 12.5 | 767.4 | 0 | 41.9 | 77.4 | 13.1 |
| 2002 |  | 340 | 200 | 5.66 | 840 | 5.46 | 445 | 7.4 | 2.8 | 7.9 | 24.8 | 1774 | 8.0 | 535.1 | 0 | 52.5 | 77.4 | 17.7 |
| 2003 |  | 503 | 247 | 2.60 | 414 | 3.76 | 400 | 7.5 | 4.1 | 9.5 | 23.9 | 1728.3 | 14.0 | 863.7 | 0 | 48.7 | 73.8 | 18.6 |
| 2004 |  | 540 | 249 | 3.17 | 472 | 4.20 | 255 | 11.2 | 2.1 | 8.6 | 23.5 | 1727 | 7.4 | 419.9 | 0 | 36.1 | 72.5 | 14.2 |
| 2005 |  | 367 | 177 | 4.85 | 540 | 4.67 | 381 | 12.8 | 2.0 | 8.0 | 16.7 | 1313.6 | 11.6 | 627.8 | 1 | 42.1 | 68.97 | 14.7 |
| 2006 |  | 356 | 166 | 6.50 | 610 | 2.19 | 202 | 10.8 | 1.4 | 6.2 | 5.2 | 478.5 | 4.6 | 280.1 | 10.9 | 28.9 | 66.84 | 12.2 |
| 2007 |  | 432 | 190 | 17.94 | 756 | 4.22 | 550 | 6.9 | 1.3 | 6.1 | 4.4 | 397.2 | 3.2 | 193.5 | 12.6 | 23.8 | 75.86 | 14.2 |
| 2008 |  | 416 | 189 | 9.03 | 469 | 4.47 | 267 | 9.5 | 0.9 | 5.1 | 2.7 | 320.4 | 1.3 | 98 | 11.5 | 12.4 | 59.94 | 9.5 |
| 2009 |  | 467 | 199 | 6.46 | 338 | 1.21 | 169 | 10.1 | 1.1 | 3.8 | 1.5 | 157.7 | 0.46 | 24.9 | 10.0 | 14.7 | 42.8 | 7.6 |
| 2010 |  | 400 | 164 | 11.55 | 371 | 14.39 | 151 | 7.9 | 1.0 | 4.8 | 1.0 | 151 | 0.19 | 10.2 | 9.2 | 15.2 | 45.8 | 9.4 |
| 2011 |  | 417 | 140 | 4.35 | 183 | 11.95 | 701 | 17.3 | 1.2 | 6.8 | 0.69 | 72.7 | 1.56 | 91.2 | 8.6 | 16.4 | 54.5 | 8.1 |
| 2012 |  | 460 | 188 | 0.74 | 276 | 7.25 | 164 | 14.9 | 1.0 | 5.0 | 0.4 | 85 | 0.9 | 60.7 | 12.1 | 14.5 | 58.3 | 7.2 |
| 2013 |  | 550 | 207 | 7.41 | 236 | -8 | 0 | 14.0 | 1.6 | 5.4 | 0.3 | 31.9 | -8 | 1.3 | 10.6 | 8.9 | 42.6 | 5.0 |
| 2014 |  | 592 | 255 | - | 87 | 8 | 0 | 13.9 | 1.5 | 8.3 | - | 16.1 | -8 | 0.4 | 8.3 | 5.1 | 47.8 | 6.0 |
| 2015 |  | 564 | 230 | - | 0 | -8 | 48 | 20.4 | 3.3 | 8.6 | - | 0 | -8 | 0.9 | 4.5 | 4.6 | 39.8 | 8.3 |
| 2016 |  | 582 | 220 | - | 0 | -8 | 0 | 26.4 | 4.6 | 32.8 | - | 0 | -8 | 3.9 | 2.5 | 2.5 | 33.4 | 7.9 |
| 2017 |  | 525 | 170 | - | 244 | -8 | 0 | 17.1 | 11.3 | 35.4 | - | 160.7 | -8 | 0 | 0.3 | 4.2 | 12.1 | 7.5 |
| 2018 |  | 554 | 139 | - | 237 | -8 | 0 | 14.6 | 8.4 | 19.5 | - | 238 | 8 | 0 | - | 3.5 | 13.6 | 9.6 |
| 2020 |  | - | - | - | 1239 | -8 | 277 | 5.9 | 4.5 | 10.6 | - | 73 | -8 | 199 | - | 13.6 | 13.3 | 10.4 |
| 2021 |  | 265 | - | - | 852 | -8 | 203 | 6.6 | 1.4 | 9.6 | - | 194 | -8 | 223 | - | 14.3 | 13.2 | 8.9 |

$1 \mathrm{Kg} / 100 \mathrm{~km}$. Sept Prime: ISS/ISN Traditional Prime Stations Only. Sept Extended: ISS/ISN/ISW/SGC All Stations.
2 Whole weight (kg) per corrected hour fished, weighted by area
3 '000 hours fished (corrected for fishing power GRT)
4 days fished
5 Corrected for fishing power (HP) [data for 1999-2010, replaced at 2011WG following recalculation at WKFLAT 2011]
6 Carhelmar survey, $\mathrm{Kg} / 100 \mathrm{~km}$ not available
7 All years updated in 2007 due to slight historical differences
8 Effort not reported in hours for this fleet, see Section 6.7.2 for more detail
9 '000s hours
Fishing power corrections are detailed in Appendix 2 of the 2000 working group report

Table 24.3a. Irish Sea plaice: NIGFS-WIBTS-Q1 indices of relative biomass trends by region in spring.

| NIGFS-WIBTS-Q1 <br> Mar (Spring) | ESTIMATED MEAN ABUNDANCE (kg/3 miles) |  |  | ESTIMATED STANDARD ERROR |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Combined | West | East | Combined | West | East |
| Year | Str 1-7 | Str 1-3 | Str 4-7 | Str 1-7 | Str 1-3 | Str 4-7 |
| 1992 | 8.35 | 5.47 | 9.20 | 3.45 | 1.96 | 4.44 |
| 1993 | 12.36 | 18.43 | 10.54 | 2.14 | 4.78 | 2.39 |
| 1994 | 9.65 | 4.47 | 11.09 | 2.43 | 1.46 | 3.12 |
| 1995 | 7.27 | 4.79 | 7.64 | 1.24 | 0.83 | 1.59 |
| 1996 | 7.29 | 12.60 | 5.70 | 1.64 | 5.71 | 1.28 |
| 1997 | 13.87 | 14.72 | 13.54 | 3.19 | 5.68 | 3.77 |
| 1998 | 10.40 | 13.32 | 9.00 | 2.73 | 7.10 | 2.84 |
| 1999 | 10.71 | 13.53 | 9.59 | 1.81 | 4.92 | 1.84 |
| 2000 | 12.92 | 26.29 | 8.88 | 4.11 | 17.00 | 1.66 |
| 2001 | 12.06 | 18.03 | 9.92 | 1.41 | 4.25 | 1.31 |
| 2002 | 15.27 | 27.95 | 11.17 | 2.53 | 8.39 | 2.14 |
| 2003 | 20.97 | 40.71 | 15.09 | 6.11 | 23.98 | 3.44 |
| 2004 | 8.55 | 5.69 | 9.40 | 1.74 | 1.21 | 2.24 |
| 2005 | 11.10 | 19.43 | 8.62 | 1.93 | 5.99 | 1.76 |
| 2006 | 7.85 | 12.14 | 6.39 | 1.39 | 4.62 | 1.16 |
| 2007 | 6.25 | 14.47 | 3.80 | 1.27 | 4.80 | 0.83 |
| 2008 | 4.46 | 5.11 | 4.57 | 0.76 | 1.23 | 0.91 |
| 2009 | 7.90 | 7.85 | 7.86 | 1.27 | 2.04 | 1.53 |
| 2010 | 19.40 | 8.77 | 17.30 | 1.86 | 2.70 | 2.28 |
| 2011 | 16.34 | 26.20 | 13.03 | 3.51 | 10.11 | 3.41 |
| 2012 | 14.22 | 21.47 | 11.05 | 2.37 | 7.48 | 2.13 |
| 2013 | 21.89 | 28.98 | 16.57 | 3.74 | 8.04 | 4.21 |
| 2014 | 11.43 | 10.96 | 9.65 | 2.04 | 4.82 | 2.22 |
| 2015 | 22.81 | 22.57 | 18.66 | 2.84 | 7.18 | 3.01 |
| 2016 | 34.52 | 30.29 | 35.77 | 7.17 | 9.95 | 8.82 |
| 2017 | 16.10 | 14.85 | 16.47 | 3.16 | 3.90 | 3.70 |
| 2018 | 19.26 | 22.86 | 18.18 | 4.11 | 10.19 | 4.39 |
| 2019 | 16.42 | 19.83 | 15.40 | 3.41 | 6.18 | 4.03 |
| 2020 | 17.69 | 12.84 | 19.13 | 3.47 | 3.79 | 4.36 |
| 2021 | 13.25 | 11.27 | 13.85 | 2.91 | 4.91 | 3.48 |

Table 24.3b. Irish Sea plaice: NIGFS-WIBTS-Q4 indices of relative biomass trends by region in autumn.

| NIGFS-WIBTS-Q4 <br> Oct (Autumn) | ESTIMATED MEAN ABUNDANCE (kg/3 miles) |  |  | ESTIMATED STANDARD ERROR |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Combined | West | East | Combined | West | East |
| Year | Str 1-7 | Str 1-3 | Str 4-7 | Str 1-7 | Str 1-3 | Str 4-7 |
| 1992 | 4.81 | 2.31 | 5.55 | 0.92 | 1.10 | 1.15 |
| 1993 | 4.48 | 2.08 | 5.20 | 1.00 | 0.87 | 1.27 |
| 1994 | 8.73 | 5.49 | 9.69 | 2.30 | 2.83 | 2.86 |
| 1995 | 4.17 | 5.50 | 3.77 | 1.13 | 2.23 | 1.31 |
| 1996 | 8.68 | 8.85 | 8.63 | 2.25 | 5.94 | 2.33 |
| 1997 | 7.93 | 5.76 | 8.58 | 2.24 | 2.59 | 2.80 |
| 1998 | 5.33 | 3.68 | 5.82 | 1.46 | 2.48 | 1.74 |
| 1999 | 5.81 | 4.30 | 6.26 | 1.67 | 3.08 | 1.97 |
| 2000 | 9.75 | 2.20 | 12.00 | 5.76 | 1.13 | 7.47 |
| 2001 | 13.85 | 2.30 | 17.30 | 6.57 | 1.67 | 8.51 |
| 2002 | 9.80 | 5.90 | 10.97 | 3.91 | 3.61 | 4.97 |
| 2003 | 18.01 | 7.52 | 21.14 | 5.84 | 4.16 | 7.48 |
| 2004 | 7.79 | 1.64 | 9.63 | 1.80 | 0.81 | 2.33 |
| 2005 | 11.35 | 3.41 | 13.72 | 4.51 | 2.18 | 5.82 |
| 2006 | 6.61 | 2.56 | 7.82 | 1.53 | 1.42 | 1.94 |
| 2007 | 7.15 | 4.07 | 8.07 | 1.41 | 2.00 | 1.73 |
| 2008 | 8.68 | 3.28 | 10.27 | 2.20 | 2.09 | 2.78 |
| 2009 | 12.44 | 4.06 | 15.01 | 2.59 | 3.12 | 3.23 |
| 2010 | 15.58 | 5.83 | 18.53 | 5.26 | 5.21 | 6.65 |
| 2011 | 14.48 | 5.39 | 15.94 | 3.55 | 2.66 | 4.55 |
| 2012 | 16.05 | 17.89 | 15.65 | 4.43 | 11.16 | 4.68 |
| 2013 | 17.90 | 13.55 | 19.09 | 4.33 | 11.27 | 4.51 |
| 2014 | 22.18 | 27.67 | 20.35 | 7.61 | 24.88 | 6.52 |
| 2015 | 18.21 | 11.15 | 20.31 | 4.39 | 8.76 | 5.06 |
| 2016 | 17.57 | 0.95 | 22.53 | 4.52 | 0.43 | 5.86 |
| 2017 | 18.55 | 2.96 | 23.20 | 4.25 | 1.59 | 5.50 |
| 2018 | 21.62 | 20.66 | 21.90 | 5.57 | 18.24 | 4.77 |
| 2019 | 16.63 | 9.50 | 18.76 | 4.06 | 6.89 | 4.86 |
| 2020 | 18.07 | 3.39 | 22.45 | 4.09 | 1.98 | 5.29 |
| 2021 | 14.34 | 4.06 | 17.41 | 3.31 | 2.08 | 4.25 |

Table 24.4. Irish Sea plaice: UK (E\&W)-BTS-Q3 biomass index (extended area). Ages in bold are those used in the assessment (ages 1-7).

| Year | Distance towed (kms) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 292.77 | 0.13 | 4.64 | 4.03 | 0.82 | 0.43 | 0.03 | 0.04 | 0.08 | 0.01 | 0.02 |
| 1994 | 218.65 | 0.33 | 4.13 | 2.48 | 1.42 | 0.28 | 0.10 | 0.03 | 0.02 | 0.03 | 0.04 |
| 1995 | 218.65 | 0.78 | 5.56 | 1.96 | 0.84 | 0.41 | 0.07 | 0.05 | 0.02 | 0.00 | 0.03 |
| 1996 | 222.36 | 0.26 | 5.79 | 2.17 | 0.53 | 0.19 | 0.20 | 0.05 | 0.02 | 0.00 | 0.02 |
| 1997 | 218.65 | 0.96 | 5.47 | 2.91 | 1.26 | 0.30 | 0.16 | 0.17 | 0.05 | 0.02 | 0.03 |
| 1998 | 218.65 | 0.56 | 4.50 | 4.26 | 1.09 | 0.38 | 0.21 | 0.08 | 0.06 | 0.01 | 0.04 |
| 1999 | 214.95 | 1.86 | 3.96 | 3.91 | 1.99 | 0.68 | 0.29 | 0.09 | 0.07 | 0.03 | 0.05 |
| 2000 | 218.65 | 1.22 | 8.74 | 2.80 | 1.47 | 1.11 | 0.47 | 0.12 | 0.09 | 0.03 | 0.04 |
| 2001 | 214.95 | 0.83 | 5.99 | 3.62 | 1.11 | 0.60 | 0.54 | 0.11 | 0.06 | 0.02 | 0.01 |
| 2002 | 214.95 | 0.23 | 6.46 | 4.94 | 2.27 | 0.88 | 0.53 | 0.48 | 0.10 | 0.04 | 0.04 |
| 2003 | 211.24 | 2.07 | 6.12 | 5.85 | 2.61 | 1.58 | 0.58 | 0.38 | 0.25 | 0.07 | 0.07 |
| 2004 | 214.95 | 1.09 | 8.07 | 5.36 | 3.94 | 1.88 | 1.15 | 0.21 | 0.19 | 0.13 | 0.10 |
| 2005 | 211.24 | 1.75 | 3.76 | 4.75 | 1.98 | 1.42 | 0.80 | 0.48 | 0.11 | 0.09 | 0.06 |
| 2006 | 214.95 | 3.56 | 5.01 | 3.45 | 2.46 | 1.10 | 0.79 | 0.36 | 0.20 | 0.02 | 0.07 |
| 2007 | 214.95 | 1.15 | 7.97 | 4.47 | 1.66 | 1.20 | 0.65 | 0.33 | 0.25 | 0.14 | 0.06 |
| 2008 | 200.12 | 1.22 | 4.68 | 5.71 | 2.03 | 1.15 | 0.82 | 0.31 | 0.12 | 0.08 | 0.05 |
| 2009 | 214.95 | 1.23 | 4.74 | 3.40 | 3.30 | 0.99 | 0.66 | 0.63 | 0.16 | 0.11 | 0.20 |
| 2010 | 211.24 | 2.01 | 6.22 | 4.31 | 2.05 | 1.44 | 0.66 | 0.54 | 0.36 | 0.20 | 0.19 |
| 2011 | 211.24 | 1.02 | 6.73 | 4.28 | 1.75 | 1.00 | 1.08 | 0.47 | 0.27 | 0.24 | 0.37 |
| 2012 | 214.95 | 1.40 | 6.52 | 6.37 | 1.71 | 1.03 | 0.47 | 0.53 | 0.30 | 0.14 | 0.42 |
| 2013 | 214.95 | 2.04 | 4.33 | 5.05 | 3.08 | 1.60 | 1.07 | 0.47 | 0.44 | 0.20 | 0.42 |
| 2014 | 214.95 | 1.56 | 7.82 | 6.85 | 3.13 | 2.16 | 0.99 | 0.77 | 0.44 | 0.20 | 0.28 |
| 2015 | 214.95 | 1.02 | 6.16 | 6.88 | 2.60 | 1.80 | 1.04 | 0.66 | 0.37 | 0.19 | 0.50 |
| 2016 | 211.24 | 0.18 | 2.91 | 5.97 | 3.95 | 2.45 | 1.61 | 0.96 | 0.74 | 0.45 | 0.58 |
| 2017 | 214.95 | 0.03 | 1.35 | 4.77 | 2.81 | 2.23 | 1.84 | 0.75 | 0.59 | 0.38 | 0.26 |
| 2018 | 214.95 | 0.36 | 1.97 | 2.75 | 2.28 | 1.51 | 1.37 | 1.24 | 0.75 | 0.56 | 0.27 |
| 2019 | 214.95 | 0.33 | 3.02 | 4.50 | 2.31 | 1.48 | 1.22 | 1.00 | 0.90 | 0.41 | 0.20 |
| 2020 | 0 | - | - | - | - | - | - | - | - | - | - |
| 2021 | 214.95 | 0.23 | 1.77 | 2.50 | 1.49 | 0.72 | 0.55 | 0.44 | 0.28 | 0.23 | 0.06 |
| Year | Distance towed (kms) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ |

Table 24.5. Irish Sea plaice: Landings number-at-age 1 to 8+ (thousands), where rows are years 1981-2021 and columns are ages 1 to 8+.

| IRISH SEA PLAICE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 12 |  |  |  |  |  |  |  |
| 19812021 |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| 22 | 1742 | 5939 | 2984 | 837 | 222 | 105 | 236 |
| 27 | 715 | 3288 | 3082 | 1358 | 330 | 137 | 213 |
| 51 | 2924 | 2494 | 3211 | 1521 | 648 | 211 | 252 |
| 41 | 3159 | 5179 | 1182 | 1054 | 459 | 299 | 252 |
| 4 | 2357 | 6152 | 3301 | 614 | 429 | 262 | 340 |
| 31 | 1652 | 5280 | 2942 | 1287 | 344 | 371 | 308 |
| 62 | 3717 | 5317 | 5252 | 1341 | 1072 | 123 | 338 |
| 46 | 2923 | 5040 | 2552 | 1400 | 750 | 316 | 405 |
| 24 | 1735 | 5945 | 2671 | 854 | 436 | 214 | 364 |
| 15 | 1019 | 2715 | 2935 | 1132 | 465 | 259 | 223 |
| 180 | 2008 | 1506 | 1929 | 1205 | 465 | 182 | 226 |
| 151 | 1958 | 3209 | 1435 | 1358 | 903 | 388 | 294 |
| 28 | 910 | 1649 | 1357 | 474 | 556 | 377 | 302 |
| 97 | 1146 | 2173 | 1309 | 644 | 318 | 245 | 263 |
| 21 | 961 | 1703 | 1936 | 764 | 318 | 138 | 157 |
| 37 | 856 | 1345 | 1196 | 943 | 370 | 128 | 135 |
| 28 | 830 | 1590 | 1513 | 1003 | 482 | 285 | 257 |
| 6 | 691 | 1739 | 1025 | 612 | 476 | 403 | 385 |
| 68 | 803 | 1505 | 1294 | 696 | 280 | 196 | 242 |
| 0 | 450 | 1174 | 1284 | 686 | 212 | 219 | 203 |
| 14 | 374 | 1138 | 1083 | 767 | 409 | 179 | 166 |
| 1 | 206 | 940 | 1482 | 842 | 539 | 318 | 170 |
| 0 | 286 | 1031 | 1314 | 707 | 415 | 253 | 222 |
| 8 | 198 | 967 | 1104 | 705 | 247 | 114 | 186 |
| 6 | 228 | 708 | 1177 | 890 | 461 | 204 | 213 |
| 5 | 180 | 620 | 550 | 684 | 346 | 220 | 218 |
| 0 | 64 | 351 | 860 | 507 | 401 | 151 | 164 |
| 1 | 99 | 386 | 389 | 409 | 215 | 141 | 119 |
| 0 | 13 | 204 | 374 | 351 | 272 | 117 | 120 |
| 0 | 7 | 75 | 271 | 306 | 193 | 160 | 115 |
| 2 | 53 | 199 | 357 | 483 | 305 | 194 | 191 |
| 0 | 8 | 150 | 292 | 301 | 367 | 218 | 226 |
| 1 | 16 | 87 | 203 | 166 | 149 | 144 | 165 |
| 3 | 6 | 65 | 165 | 160 | 143 | 70 | 158 |
| 0 | 1 | 43 | 93 | 185 | 210 | 149 | 349 |
| 14 | 14 | 58 | 162 | 224 | 346 | 180 | 482 |
| 0 | 4 | 24 | 145 | 206 | 241 | 209 | 520 |
| 0 | 6 | 84 | 110 | 201 | 178 | 151 | 358 |
| 0 | 11 | 53 | 145 | 273 | 219 | 187 | 356 |
| 2 | 17 | 24 | 118 | 192 | 168 | 150 | 287 |
| 0 | 30 | 80 | 146 | 154 | 106 | 127 | 199 |

Table 24.6. Irish Sea plaice: Landings weight-at-age 1 to 8+(kg), where rows are years 1981-2021 and columns are ages 1 to 8+

IRISH SEA PLAICE
13
19812021
18
1

| 0.069 | 0.176 | 0.267 | 0.376 | 0.512 | 0.592 | 0.678 | 1.085 |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.201 | 0.274 | 0.284 | 0.348 | 0.421 | 0.545 | 0.650 | 0.889 |
| 0.232 | 0.261 | 0.290 | 0.319 | 0.368 | 0.426 | 0.484 | 0.699 |
| 0.260 | 0.290 | 0.330 | 0.380 | 0.470 | 0.560 | 0.660 | 0.964 |
| 0.290 | 0.310 | 0.340 | 0.390 | 0.470 | 0.540 | 0.630 | 0.851 |
| 0.270 | 0.280 | 0.340 | 0.420 | 0.500 | 0.540 | 0.630 | 0.980 |
| 0.260 | 0.290 | 0.315 | 0.370 | 0.440 | 0.520 | 0.610 | 0.916 |
| 0.230 | 0.260 | 0.300 | 0.370 | 0.460 | 0.550 | 0.680 | 1.243 |
| 0.227 | 0.272 | 0.321 | 0.374 | 0.430 | 0.491 | 0.555 | 0.761 |
| 0.200 | 0.257 | 0.316 | 0.376 | 0.439 | 0.504 | 0.570 | 0.747 |
| 0.247 | 0.267 | 0.295 | 0.332 | 0.377 | 0.431 | 0.494 | 0.652 |
| 0.169 | 0.218 | 0.274 | 0.337 | 0.407 | 0.484 | 0.568 | 0.799 |
| 0.260 | 0.270 | 0.292 | 0.328 | 0.375 | 0.436 | 0.508 | 0.690 |
| 0.156 | 0.207 | 0.268 | 0.338 | 0.416 | 0.504 | 0.600 | 0.816 |
| 0.189 | 0.224 | 0.262 | 0.329 | 0.353 | 0.406 | 0.461 | 0.699 |
| 0.204 | 0.223 | 0.270 | 0.333 | 0.398 | 0.493 | 0.584 | 0.837 |
| 0.205 | 0.233 | 0.241 | 0.286 | 0.354 | 0.410 | 0.510 | 0.620 |
| 0.185 | 0.226 | 0.249 | 0.316 | 0.353 | 0.410 | 0.468 | 0.655 |
| 0.205 | 0.236 | 0.250 | 0.300 | 0.375 | 0.457 | 0.483 | 0.615 |
| 0.000 | 0.259 | 0.270 | 0.307 | 0.337 | 0.429 | 0.437 | 0.623 |
| 0.232 | 0.233 | 0.271 | 0.334 | 0.396 | 0.439 | 0.571 | 0.764 |
| 0.228 | 0.271 | 0.267 | 0.308 | 0.386 | 0.476 | 0.518 | 0.673 |
| 0.000 | 0.235 | 0.289 | 0.335 | 0.383 | 0.458 | 0.567 | 0.678 |
| 0.214 | 0.239 | 0.258 | 0.297 | 0.347 | 0.416 | 0.543 | 0.571 |
| 0.235 | 0.245 | 0.265 | 0.292 | 0.322 | 0.394 | 0.441 | 0.632 |
| 0.200 | 0.256 | 0.265 | 0.282 | 0.321 | 0.378 | 0.425 | 0.568 |
| 0.000 | 0.280 | 0.266 | 0.281 | 0.320 | 0.371 | 0.416 | 0.481 |
| 0.246 | 0.228 | 0.257 | 0.281 | 0.311 | 0.364 | 0.431 | 0.553 |
| 0.000 | 0.257 | 0.256 | 0.265 | 0.305 | 0.330 | 0.395 | 0.482 |
| 0.000 | 0.260 | 0.265 | 0.282 | 0.301 | 0.356 | 0.392 | 0.492 |
| 0.236 | 0.251 | 0.257 | 0.283 | 0.298 | 0.354 | 0.404 | 0.513 |
| 0.117 | 0.259 | 0.254 | 0.281 | 0.299 | 0.318 | 0.345 | 0.430 |
| 0.249 | 0.245 | 0.249 | 0.267 | 0.297 | 0.330 | 0.386 | 0.417 |
| 0.181 | 0.250 | 0.282 | 0.300 | 0.336 | 0.373 | 0.457 | 0.492 |
| NA | 0.183 | 0.264 | 0.287 | 0.299 | 0.340 | 0.403 | 0.617 |
| 0.113 | 0.149 | 0.229 | 0.318 | 0.422 | 0.362 | 0.433 | 0.660 |
| 0.166 | 0.222 | 0.273 | 0.345 | 0.370 | 0.405 | 0.442 | 0.505 |
| 0.000 | 0.292 | 0.327 | 0.353 | 0.345 | 0.398 | 0.399 | 0.465 |
| 0.108 | 0.251 | 0.270 | 0.283 | 0.288 | 0.350 | 0.379 | 0.509 |
| 0.107 | 0.130 | 0.190 | 0.280 | 0.331 | 0.360 | 0.363 | 0.390 |
| 0 | 0.253 | 0.281 | 0.295 | 0.299 | 0.342 | 0.343 | 0.388 |
| 0 |  |  |  |  |  |  |  |
| 0 |  |  |  |  |  |  |  |

Table 24.7. Irish Sea plaice: Discards weight-at-age 1 to 8+(kg), where rows are years 1981-2021 and columns are ages 1

## to $8+$.

| IRISH SEA PLAICE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 123 |  |  |  |  |  |  |  |
| 19812021 |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.087 | 0.105 | 0.130 | 0.153 | 0.170 | 0.231 | 0.318 | 0.211 |
| 0.057 | 0.115 | 0.145 | 0.164 | 0.211 | 0.290 | 0.238 | 0.210 |
| 0.099 | 0.117 | 0.134 | 0.179 | 0.178 | 0.277 | 0.644 | 0.356 |
| 0.141 | 0.113 | 0.141 | 0.145 | 0.162 | 0.210 | 0.274 | 0.077 |
| 0.044 | 0.081 | 0.113 | 0.140 | 0.150 | 0.205 | 0.219 | 0.243 |
| 0.096 | 0.097 | 0.116 | 0.135 | 0.151 | 0.173 | 0.217 | 0.170 |
| 0.033 | 0.080 | 0.119 | 0.147 | 0.165 | 0.196 | 0.232 | 0.276 |
| 0.083 | 0.101 | 0.138 | 0.183 | 0.201 | 0.140 | 0.194 | 0.225 |
| 0.077 | 0.098 | 0.116 | 0.141 | 0.157 | 0.168 | 0.164 | 0.176 |
| 0.026 | 0.038 | 0.081 | 0.119 | 0.162 | 0.200 | 0.157 | 0.182 |
| 0.064 | 0.069 | 0.094 | 0.116 | 0.144 | 0.157 | 0.181 | 0.181 |
| 0.056 | 0.067 | 0.084 | 0.120 | 0.128 | 0.150 | 0.152 | 0.153 |
| 0.088 | 0.059 | 0.079 | 0.101 | 0.095 | 0.126 | 0.152 | 0.136 |
| 0.136 | 0.103 | 0.109 | 0.120 | 0.146 | 0.161 | 0.155 | 0.170 |
| 0.093 | 0.080 | 0.118 | 0. 124 | 0.128 | 0. 153 | 0. 137 | 0.157 |
| 0.022 | 0.053 | 0.075 | 0.109 | 0.142 | 0.143 | 0.146 | 0.202 |
| 0.054 | 0.062 | 0.082 | 0.104 | 0.127 | 0.136 | 0.167 | 0.149 |
| 0.513 | 0.081 | 0.103 | 0.121 | 0.133 | 0.150 | 0.157 | 0.151 |
| 0.040 | 0.095 | 0.121 | 0.1387 | 0.167 | 0.172 | 0.204 | 0.258 |

Table 24.8. Irish Sea plaice: New stock weights-at-age modified to include discard element (kg), where rows are years 1981-2021 and columns are ages 1 to 8+.

| IRISH SEA PLAICE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14 |  |  |  |  |  |  |  |
| 19812021 |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| 0.087 | 0.124 | 0.190 | 0.351 | 0.509 | 0.592 | 0.678 | 1.085 |
| 0.091 | 0.141 | 0.210 | 0.327 | 0.418 | 0.545 | 0.650 | 0.889 |
| 0.097 | 0.173 | 0.231 | 0.303 | 0.366 | 0.426 | 0.484 | 0.699 |
| 0.100 | 0.196 | 0.275 | 0.362 | 0.467 | 0.560 | 0.660 | 0.964 |
| 0.089 | 0.203 | 0.293 | 0.374 | 0.468 | 0.540 | 0.630 | 0.851 |
| 0.098 | 0.171 | 0.292 | 0.401 | 0.497 | 0.540 | 0.630 | 0.980 |
| 0.102 | 0.208 | 0.266 | 0.353 | 0.437 | 0.519 | 0.610 | 0.916 |
| 0.104 | 0.171 | 0.250 | 0.351 | 0.456 | 0.549 | 0.680 | 1.243 |
| 0.100 | 0.183 | 0.261 | 0.352 | 0.425 | 0.490 | 0.555 | 0.761 |
| 0.090 | 0.172 | 0.253 | 0.349 | 0.431 | 0.502 | 0.570 | 0.747 |
| 0.140 | 0.165 | 0.230 | 0.305 | 0.369 | 0.429 | 0.494 | 0.652 |
| 0.106 | 0.159 | 0.209 | 0.302 | 0.395 | 0.481 | 0.568 | 0.799 |
| 0.097 | 0.141 | 0.209 | 0.291 | 0.363 | 0.434 | 0.508 | 0.690 |
| 0.101 | 0.134 | 0.193 | 0.299 | 0.400 | 0.501 | 0.600 | 0.816 |
| 0.091 | 0.138 | 0.184 | 0.289 | 0.340 | 0.404 | 0.461 | 0.699 |
| 0.091 | 0.130 | 0.181 | 0.286 | 0.377 | 0.488 | 0.583 | 0.837 |
| 0.091 | 0.118 | 0.168 | 0.247 | 0.335 | 0.406 | 0.509 | 0.620 |
| 0.088 | 0.116 | 0.148 | 0.223 | 0.305 | 0.399 | 0.466 | 0.655 |
| 0.100 | 0.125 | 0.150 | 0.216 | 0.321 | 0.444 | 0.480 | 0.615 |
| NA | 0.121 | 0.157 | 0.222 | 0.300 | 0.420 | 0.436 | 0.623 |
| 0.091 | 0.119 | 0.161 | 0.239 | 0.352 | 0.431 | 0.569 | 0.764 |
| 0.088 | 0.114 | 0.161 | 0.228 | 0.347 | 0.467 | 0.517 | 0.673 |
| NA | 0.115 | 0.165 | 0.234 | 0.335 | 0.448 | 0.566 | 0.678 |
| 0.070 | 0.131 | 0.169 | 0.217 | 0.304 | 0.407 | 0.540 | 0.570 |
| 0.103 | 0.127 | 0.161 | 0.238 | 0.234 | 0.377 | 0.454 | 0.602 |
| 0.141 | 0.122 | 0.162 | 0.175 | 0.256 | 0.323 | 0.417 | 0.564 |
| 0.044 | 0.084 | 0.123 | 0.167 | 0.209 | 0.290 | 0.335 | 0.377 |
| 0.096 | 0.100 | 0.131 | 0.168 | 0.204 | 0.279 | 0.397 | 0.285 |
| 0.033 | 0.081 | 0.125 | 0.173 | 0.213 | 0.266 | 0.333 | 0.413 |
| 0.083 | 0.101 | 0.140 | 0.191 | 0.211 | 0.190 | 0.226 | 0.290 |
| 0.078 | 0.104 | 0.137 | 0.182 | 0.221 | 0.271 | 0.334 | 0.364 |
| 0.026 | 0.038 | 0.088 | 0.142 | 0.199 | 0.246 | 0.232 | 0.294 |
| 0.065 | 0.071 | 0.098 | 0.133 | 0.185 | 0.240 | 0.292 | 0.363 |
| 0.056 | 0.068 | 0.089 | 0.135 | 0.153 | 0.194 | 0.214 | 0.296 |
| 0.088 | 0.060 | 0.083 | 0.115 | 0.130 | 0.163 | 0.269 | 0.515 |
| 0.133 | 0.105 | 0.117 | 0.152 | 0.240 | 0.259 | 0.307 | 0.522 |
| 0.093 | 0. 081 | 0.121 | 0.145 | 0. 163 | 0.198 | 0.223 | 0.303 |
| 0.022 | 0.054 | 0.098 | 0.138 | 0.199 | 0.253 | 0.269 | 0.39 |
| 0.054 | 0.062 | 0.088 | 0.127 | 0.180 | 0.218 | 0.304 | 0.427 |
| 0.063 | 0.084 | 0.106 | 0.151 | 0.198 | 0.240 | 0.269 | 0.298 |
| 0.040 | 0.01 | 0.131 | 0.170 | 0.227 | 0.236 | 0.289 | 0.381 |

Table 24.9. Irish Sea plaice: Estimated landed numbers-at-age (thousands).

| year\age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 22 | 1742 | 5939 | 2984 | 837 | 222 | 105 | 236 | 12087 |
| 1982 | 27 | 715 | 3288 | 3082 | 1358 | 330 | 137 | 213 | 9150 |
| 1983 | 51 | 2924 | 2494 | 3211 | 1521 | 648 | 211 | 252 | 11312 |
| 1984 | 41 | 3159 | 5179 | 1182 | 1054 | 459 | 299 | 252 | 11625 |
| 1985 | 4 | 2357 | 6152 | 3301 | 614 | 429 | 262 | 340 | 13459 |
| 1986 | 31 | 1652 | 5280 | 2942 | 1287 | 344 | 371 | 308 | 12215 |
| 1987 | 62 | 3717 | 5317 | 5252 | 1341 | 1072 | 123 | 338 | 17222 |
| 1988 | 46 | 2923 | 5040 | 2552 | 1400 | 750 | 316 | 405 | 13432 |
| 1989 | 24 | 1735 | 5945 | 2671 | 854 | 436 | 214 | 364 | 12243 |
| 1990 | 15 | 1019 | 2715 | 2935 | 1132 | 465 | 259 | 223 | 8763 |
| 1991 | 180 | 2008 | 1506 | 1929 | 1205 | 465 | 182 | 226 | 7701 |
| 1992 | 151 | 1958 | 3209 | 1435 | 1358 | 903 | 388 | 294 | 9696 |
| 1993 | 28 | 910 | 1649 | 1357 | 474 | 556 | 377 | 302 | 5653 |
| 1994 | 97 | 1146 | 2173 | 1309 | 644 | 318 | 245 | 263 | 6195 |
| 1995 | 21 | 961 | 1703 | 1936 | 764 | 318 | 138 | 157 | 5998 |
| 1996 | 37 | 856 | 1345 | 1196 | 943 | 370 | 128 | 135 | 5011 |
| 1997 | 28 | 830 | 1590 | 1513 | 1003 | 482 | 285 | 257 | 5988 |
| 1998 | 6 | 691 | 1739 | 1025 | 612 | 476 | 403 | 385 | 5336 |
| 1999 | 68 | 803 | 1505 | 1294 | 696 | 280 | 196 | 242 | 5083 |
| 2000 | 0 | 450 | 1174 | 1284 | 686 | 212 | 219 | 203 | 4228 |
| 2001 | 14 | 374 | 1138 | 1083 | 767 | 409 | 178 | 166 | 4130 |
| 2002 | 1 | 206 | 940 | 1482 | 842 | 539 | 318 | 170 | 4497 |
| 2003 | 0 | 286 | 1031 | 1314 | 707 | 415 | 253 | 222 | 4227 |
| 2004 | 8 | 198 | 967 | 1104 | 705 | 247 | 114 | 186 | 3529 |
| 2005 | 6 | 228 | 708 | 1177 | 890 | 461 | 204 | 213 | 3888 |
| 2006 | 5 | 180 | 620 | 550 | 684 | 346 | 220 | 218 | 2823 |
| 2007 | 0 | 64 | 351 | 860 | 507 | 401 | 151 | 164 | 2497 |
| 2008 | 1 | 99 | 386 | 389 | 409 | 215 | 141 | 119 | 1757 |
| 2009 | 0 | 13 | 204 | 374 | 351 | 272 | 117 | 120 | 1451 |
| 2010 | 0 | 7 | 75 | 271 | 306 | 193 | 160 | 115 | 1127 |
| 2011 | 2 | 53 | 199 | 357 | 483 | 305 | 194 | 191 | 1785 |
| 2012 | 0 | 8 | 150 | 292 | 301 | 367 | 218 | 226 | 1561 |
| 2013 | 1 | 16 | 87 | 203 | 166 | 149 | 144 | 165 | 931 |
| 2014 | 3 | 6 | 65 | 165 | 160 | 143 | 70 | 158 | 772 |
| 2015 | 0 | 1 | 43 | 93 | 185 | 210 | 149 | 349 | 1030 |
| 2016 | 14 | 14 | 58 | 162 | 224 | 346 | 180 | 482 | 1479 |
| 2017 | 0 | 4 | 24 | 145 | 206 | 241 | 209 | 519 | 1348 |
| 2018 | 0 | 6 | 84 | 109 | 201 | 178 | 151 | 358 | 1087 |
| 2019 | 0 | 11 | 53 | 145 | 273 | 219 | 187 | 356 | 1245 |
| 2020 | 2 | 17 | 24 | 118 | 192 | 168 | 150 | 287 | 959 |
| 2021 | 0 | 30 | 80 | 146 | 154 | 106 | 127 | 199 | 842 |

Table 24.10. Irish Sea plaice: Estimated discarded numbers-at-age (thousands). All discards are included (dead and alive portions).

| year\age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 451 | 4589 | 7613 | 377 | 7 | 0 | 0 | 0 | 13037 |
| 1982 | 765 | 2570 | 3062 | 375 | 14 | 0 | 0 | 0 | 6786 |
| 1983 | 724 | 3771 | 1457 | 346 | 18 | 1 | 0 | 0 | 6316 |
| 1984 | 532 | 3218 | 1970 | 102 | 11 | 1 | 0 | 0 | 5834 |
| 1985 | 508 | 2572 | 1781 | 232 | 5 | 1 | 0 | 0 | 5098 |
| 1986 | 495 | 2707 | 1572 | 228 | 12 | 1 | 0 | 0 | 5015 |
| 1987 | 668 | 2962 | 1917 | 446 | 14 | 2 | 0 | 0 | 6010 |
| 1988 | 360 | 3903 | 2081 | 249 | 21 | 2 | 0 | 0 | 6615 |
| 1989 | 240 | 1987 | 2710 | 290 | 17 | 2 | 0 | 0 | 5246 |
| 1990 | 604 | 1278 | 1398 | 403 | 34 | 3 | 0 | 0 | 3719 |
| 1991 | 364 | 3363 | 980 | 348 | 50 | 4 | 0 | 0 | 5109 |
| 1992 | 528 | 2124 | 2661 | 342 | 75 | 9 | 1 | 0 | 5740 |
| 1993 | 460 | 3187 | 1726 | 358 | 29 | 6 | 1 | 0 | 5767 |
| 1994 | 406 | 2849 | 2606 | 353 | 45 | 4 | 0 | 0 | 6265 |
| 1995 | 507 | 2502 | 2423 | 561 | 59 | 4 | 0 | 0 | 6057 |
| 1996 | 1205 | 3086 | 2329 | 417 | 94 | 7 | 0 | 0 | 7138 |
| 1997 | 935 | 7406 | 3079 | 619 | 116 | 11 | 1 | 0 | 12166 |
| 1998 | 686 | 6642 | 9665 | 1364 | 215 | 31 | 6 | 0 | 18609 |
| 1999 | 582 | 4459 | 7451 | 1734 | 247 | 18 | 4 | 0 | 14495 |
| 2000 | 0 | 3763 | 4922 | 1558 | 193 | 10 | 3 | 0 | 10449 |
| 2001 | 513 | 2934 | 4078 | 1201 | 186 | 16 | 1 | 0 | 8931 |
| 2002 | 490 | 3399 | 3168 | 1558 | 188 | 21 | 1 | 0 | 8825 |
| 2003 | 0 | 3281 | 3685 | 1623 | 204 | 19 | 1 | 0 | 8813 |
| 2004 | 85 | 1381 | 3570 | 1679 | 324 | 19 | 1 | 0 | 7059 |
| 2005 | 198 | 2844 | 2793 | 1096 | 1392 | 78 | 14 | 26 | 8441 |
| 2006 | 854 | 2775 | 2964 | 1968 | 479 | 170 | 12 | 2 | 9224 |
| 2007 | 837 | 4704 | 4892 | 3568 | 947 | 381 | 104 | 127 | 15560 |
| 2008 | 831 | 4393 | 3188 | 1354 | 837 | 171 | 27 | 278 | 11079 |
| 2009 | 56 | 2862 | 4318 | 1318 | 677 | 251 | 71 | 60 | 9613 |
| 2010 | 980 | 4066 | 4113 | 3254 | 2853 | 638 | 836 | 359 | 17099 |
| 2011 | 540 | 1344 | 1134 | 888 | 589 | 245 | 79 | 151 | 4970 |
| 2012 | 219 | 4415 | 3492 | 1755 | 800 | 567 | 329 | 274 | 11851 |
| 2013 | 238 | 1610 | 3066 | 1633 | 450 | 163 | 122 | 49 | 7331 |
| 2014 | 1027 | 1886 | 2710 | 1843 | 1149 | 591 | 274 | 218 | 9697 |
| 2015 | 18 | 1348 | 1659 | 1104 | 896 | 997 | 170 | 93 | 6285 |
| 2016 | 101 | 300 | 858 | 831 | 430 | 364 | 149 | 189 | 3222 |
| 2017 | 45 | 529 | 1057 | 1376 | 1198 | 1118 | 530 | 723 | 6576 |
| 2018 | 321 | 1464 | 823 | 814 | 524 | 235 | 159 | 143 | 4482 |
| 2019 | 167 | 2147 | 1729 | 990 | 549 | 352 | 103 | 105 | 6142 |
| 2020 | 9 | 289 | 511 | 501 | 395 | 227 | 128 | 177 | 2237 |
| 2021 | 186 | 951 | 1112 | 577 | 184 | 175 | 82 | 10 | 3278 |

Table 24.11. Irish Sea plaice: Estimated population numbers-at-age (thousands).

| year\agege | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 17044 | 19824 | 17536 | 7247 | 2022 | 694 | 329 | 765 | 65460 |
| 1982 | 22045 | 12863 | 13265 | 7693 | 3283 | 1011 | 393 | 655 | 61207 |
| 1983 | 23466 | 20215 | 8030 | 6385 | 3326 | 1661 | 568 | 635 | 64287 |
| 1984 | 22715 | 20854 | 14122 | 3416 | 2747 | 1506 | 882 | 691 | 66934 |
| 1985 | 21205 | 20162 | 14440 | 7173 | 1562 | 1418 | 817 | 927 | 67704 |
| 1986 | 21558 | 17818 | 14600 | 6979 | 3456 | 790 | 823 | 995 | 67019 |
| 1987 | 20833 | 19476 | 12300 | 7445 | 3122 | 1795 | 406 | 1032 | 66408 |
| 1988 | 15812 | 19946 | 13071 | 5348 | 2944 | 1391 | 857 | 795 | 60165 |
| 1989 | 12910 | 13498 | 14450 | 5954 | 2177 | 1277 | 644 | 858 | 51767 |
| 1990 | 15754 | 9495 | 9170 | 7424 | 2738 | 1066 | 665 | 782 | 47094 |
| 1991 | 16149 | 14163 | 5696 | 4552 | 3645 | 1365 | 552 | 778 | 46900 |
| 1992 | 17592 | 12822 | 9531 | 2505 | 2095 | 1996 | 748 | 739 | 48028 |
| 1993 | 15873 | 16302 | 7677 | 4114 | 869 | 861 | 1061 | 755 | 47511 |
| 1994 | 15159 | 12373 | 11548 | 3522 | 1642 | 491 | 431 | 924 | 46091 |
| 1995 | 17792 | 10973 | 7597 | 5505 | 1533 | 778 | 275 | 703 | 45155 |
| 1996 | 21761 | 13662 | 6861 | 3516 | 2863 | 922 | 408 | 590 | 50582 |
| 1997 | 22635 | 17734 | 9717 | 3895 | 2070 | 1792 | 661 | 686 | 59190 |
| 1998 | 19868 | 20706 | 12001 | 4776 | 2207 | 1236 | 1069 | 898 | 62761 |
| 1999 | 19050 | 16969 | 15272 | 6356 | 2548 | 1301 | 839 | 1185 | 63519 |
| 2000 | 23845 | 14769 | 11762 | 9235 | 3608 | 1431 | 977 | 1311 | 66937 |
| 2001 | 24174 | 18810 | 10518 | 6915 | 5695 | 2012 | 940 | 1511 | 70575 |
| 2002 | 24688 | 20704 | 14967 | 7069 | 4721 | 4109 | 1488 | 1684 | 79429 |
| 2003 | 22155 | 22016 | 16429 | 10947 | 4423 | 3338 | 2874 | 2257 | 84438 |
| 2004 | 20893 | 17612 | 17215 | 11422 | 7267 | 2574 | 2223 | 3231 | 82437 |
| 2005 | 18137 | 18437 | 12956 | 10751 | 7351 | 4635 | 1773 | 3530 | 77569 |
| 2006 | 22440 | 15197 | 14568 | 8524 | 6309 | 4323 | 2874 | 3375 | 77611 |
| 2007 | 26127 | 18466 | 11937 | 10472 | 5455 | 3707 | 2995 | 4009 | 83167 |
| 2008 | 21325 | 22970 | 13076 | 8323 | 7325 | 3396 | 2219 | 4671 | 83305 |
| 2009 | 17967 | 16559 | 18247 | 9051 | 6224 | 5899 | 2450 | 4764 | 81162 |
| 2010 | 23232 | 16045 | 12769 | 13115 | 7279 | 5110 | 4969 | 5540 | 88060 |
| 2011 | 26850 | 17655 | 11534 | 8233 | 9089 | 5325 | 3783 | 7521 | 89991 |
| 2012 | 24072 | 24546 | 14183 | 9408 | 6040 | 6875 | 4330 | 8589 | 98044 |
| 2013 | 23785 | 20492 | 19435 | 11882 | 7990 | 5101 | 5501 | 9365 | 103553 |
| 2014 | 27392 | 21906 | 17564 | 15190 | 9934 | 6985 | 4744 | 11470 | 115185 |
| 2015 | 17302 | 23508 | 17759 | 13808 | 11427 | 8718 | 5559 | 13213 | 111294 |
| 2016 | 14284 | 15389 | 18813 | 15642 | 11653 | 9826 | 7830 | 15267 | 108704 |
| 2017 | 10582 | 13312 | 13344 | 15115 | 13665 | 9670 | 8190 | 17888 | 101763 |
| 2018 | 12406 | 11029 | 11440 | 10741 | 11593 | 10897 | 8006 | 17693 | 93805 |
| 2019 | 11245 | 12400 | 10196 | 9472 | 8805 | 9493 | 9079 | 17401 | 88092 |
| 2020 | 8749 | 9300 | 8765 | 7457 | 7219 | 6636 | 7814 | 19036 | 74976 |
| 2021 | 9261 | 8610 | 8361 | 6628 | 5536 | 5410 | 4963 | 18552 | 67321 |

Table 24.12. Irish Sea plaice: Estimated fishing mortality-at-age.

| year\age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | $F_{\text {bar }}$ (3-6) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.020 | 0.270 | 0.642 | 0.681 | 0.574 | 0.477 | 0.423 | 0.423 | 0.594 |
| 1982 | 0.020 | 0.265 | 0.630 | 0.675 | 0.572 | 0.480 | 0.426 | 0.426 | 0.589 |
| 1983 | 0.021 | 0.282 | 0.673 | 0.727 | 0.622 | 0.527 | 0.468 | 0.468 | 0.637 |
| 1984 | 0.019 | 0.250 | 0.593 | 0.646 | 0.558 | 0.479 | 0.428 | 0.428 | 0.569 |
| 1985 | 0.018 | 0.242 | 0.575 | 0.633 | 0.551 | 0.478 | 0.428 | 0.428 | 0.560 |
| 1986 | 0.019 | 0.251 | 0.596 | 0.663 | 0.583 | 0.513 | 0.456 | 0.456 | 0.589 |
| 1987 | 0.022 | 0.295 | 0.697 | 0.780 | 0.688 | 0.605 | 0.531 | 0.531 | 0.692 |
| 1988 | 0.022 | 0.291 | 0.681 | 0.767 | 0.687 | 0.610 | 0.539 | 0.539 | 0.686 |
| 1989 | 0.020 | 0.260 | 0.597 | 0.668 | 0.600 | 0.535 | 0.475 | 0.475 | 0.600 |
| 1990 | 0.021 | 0.265 | 0.596 | 0.658 | 0.591 | 0.523 | 0.464 | 0.464 | 0.592 |
| 1991 | 0.022 | 0.277 | 0.613 | 0.672 | 0.599 | 0.525 | 0.462 | 0.462 | 0.602 |
| 1992 | 0.027 | 0.335 | 0.738 | 0.819 | 0.738 | 0.642 | 0.562 | 0.562 | 0.734 |
| 1993 | 0.024 | 0.298 | 0.651 | 0.724 | 0.665 | 0.584 | 0.510 | 0.510 | 0.656 |
| 1994 | 0.025 | 0.302 | 0.645 | 0.704 | 0.643 | 0.562 | 0.491 | 0.491 | 0.638 |
| 1995 | 0.024 | 0.291 | 0.608 | 0.644 | 0.576 | 0.495 | 0.430 | 0.430 | 0.581 |
| 1996 | 0.024 | 0.281 | 0.572 | 0.588 | 0.511 | 0.433 | 0.374 | 0.374 | 0.526 |
| 1997 | 0.025 | 0.286 | 0.575 | 0.583 | 0.501 | 0.421 | 0.363 | 0.363 | 0.520 |
| 1998 | 0.024 | 0.282 | 0.568 | 0.571 | 0.486 | 0.406 | 0.347 | 0.347 | 0.508 |
| 1999 | 0.020 | 0.227 | 0.452 | 0.451 | 0.380 | 0.311 | 0.259 | 0.259 | 0.398 |
| 2000 | 0.017 | 0.191 | 0.378 | 0.376 | 0.315 | 0.255 | 0.207 | 0.207 | 0.331 |
| 2001 | 0.015 | 0.164 | 0.326 | 0.328 | 0.275 | 0.220 | 0.174 | 0.174 | 0.287 |
| 2002 | 0.012 | 0.139 | 0.275 | 0.281 | 0.237 | 0.186 | 0.143 | 0.143 | 0.245 |
| 2003 | 0.010 | 0.116 | 0.227 | 0.233 | 0.197 | 0.152 | 0.112 | 0.112 | 0.202 |
| 2004 | 0.008 | 0.089 | 0.173 | 0.178 | 0.150 | 0.114 | 0.081 | 0.081 | 0.154 |
| 2005 | 0.011 | 0.118 | 0.224 | 0.227 | 0.191 | 0.143 | 0.098 | 0.098 | 0.196 |
| 2006 | 0.013 | 0.137 | 0.248 | 0.246 | 0.204 | 0.149 | 0.098 | 0.098 | 0.212 |
| 2007 | 0.015 | 0.154 | 0.272 | 0.266 | 0.219 | 0.159 | 0.102 | 0.102 | 0.229 |
| 2008 | 0.012 | 0.123 | 0.212 | 0.206 | 0.170 | 0.124 | 0.079 | 0.079 | 0.178 |
| 2009 | 0.009 | 0.091 | 0.157 | 0.153 | 0.129 | 0.096 | 0.061 | 0.061 | 0.134 |
| 2010 | 0.013 | 0.131 | 0.222 | 0.218 | 0.185 | 0.139 | 0.087 | 0.087 | 0.191 |
| 2011 | 0.009 | 0.086 | 0.144 | 0.143 | 0.123 | 0.095 | 0.060 | 0.060 | 0.126 |
| 2012 | 0.009 | 0.085 | 0.143 | 0.143 | 0.125 | 0.098 | 0.062 | 0.062 | 0.127 |
| 2013 | 0.006 | 0.058 | 0.096 | 0.096 | 0.085 | 0.068 | 0.043 | 0.043 | 0.086 |
| 2014 | 0.006 | 0.059 | 0.097 | 0.099 | 0.089 | 0.073 | 0.046 | 0.046 | 0.089 |
| 2015 | 0.004 | 0.037 | 0.062 | 0.066 | 0.062 | 0.054 | 0.035 | 0.035 | 0.061 |
| 2016 | 0.004 | 0.033 | 0.055 | 0.058 | 0.055 | 0.049 | 0.032 | 0.032 | 0.054 |
| 2017 | 0.005 | 0.042 | 0.068 | 0.072 | 0.068 | 0.059 | 0.038 | 0.038 | 0.067 |
| 2018 | 0.006 | 0.049 | 0.076 | 0.075 | 0.067 | 0.055 | 0.035 | 0.035 | 0.068 |
| 2019 | 0.006 | 0.055 | 0.083 | 0.082 | 0.071 | 0.057 | 0.035 | 0.035 | 0.073 |
| 2020 | 0.004 | 0.038 | 0.058 | 0.057 | 0.050 | 0.040 | 0.025 | 0.025 | 0.051 |
| 2021 | 0.006 | 0.051 | 0.077 | 0.074 | 0.063 | 0.049 | 0.029 | 0.029 | 0.066 |

Table 24.13. Irish Sea plaice: SAM stock assessment summary ( $\pm 2$ standard deviation uncertainty). Recruitment (000s), spawning-stock biomass (SSB, tonnes), mean fishing mortality (Fbar) for ages 3-6, total stock biomass (TSB, tonnes) and dead catch tonnage (the sum of landings and $60 \%$ of discards).

| Year | Recruitment <br> (thousands) |  |  | SSB (t) |  |  | Fbar (3-6) |  |  | TSB (t) |  |  | Dead catch ( t ) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High |
| 1981 | 11478 | 17044 | 25310 | 5650 | 7057 | 8814 | 0.452 | 0.594 | 0.78 | 10469 | 12830 | 15723 | 3177 | 4384 | 6049 |
| 1982 | 15679 | 22045 | 30995 | 5521 | 6792 | 8356 | 0.459 | 0.589 | 0.757 | 10369 | 12476 | 15010 | 3104 | 4069 | 5336 |
| 1983 | 16814 | 23466 | 32751 | 5030 | 6092 | 7378 | 0.497 | 0.637 | 0.817 | 10721 | 12917 | 15563 | 3045 | 3908 | 5017 |
| 1984 | 16365 | 22715 | 31529 | 6322 | 7676 | 9319 | 0.444 | 0.569 | 0.729 | 13051 | 15784 | 19090 | 3522 | 4550 | 5878 |
| 1985 | 15322 | 21205 | 29348 | 6924 | 8417 | 10232 | 0.438 | 0.56 | 0.716 | 13655 | 16510 | 19961 | 3873 | 5024 | 6516 |
| 1986 | 15547 | 21558 | 29892 | 7419 | 9019 | 10964 | 0.463 | 0.589 | 0.749 | 13906 | 16684 | 20018 | 4165 | 5382 | 6954 |
| 1987 | 14943 | 20833 | 29044 | 6984 | 8431 | 10176 | 0.546 | 0.692 | 0.878 | 13703 | 16421 | 19678 | 4411 | 5667 | 7281 |
| 1988 | 11429 | 15812 | 21876 | 6533 | 7901 | 9556 | 0.54 | 0.686 | 0.872 | 12263 | 14661 | 17528 | 4016 | 5133 | 6560 |
| 1989 | 9099 | 12910 | 18316 | 5806 | 7055 | 8573 | 0.471 | 0.6 | 0.764 | 10713 | 12894 | 15519 | 3344 | 4338 | 5627 |
| 1990 | 11449 | 15754 | 21678 | 5293 | 6436 | 7826 | 0.466 | 0.592 | 0.752 | 9300 | 11115 | 13284 | 2913 | 3755 | 4842 |
| 1991 | 11815 | 16149 | 22074 | 4247 | 5125 | 6185 | 0.475 | 0.602 | 0.762 | 9036 | 10775 | 12849 | 2371 | 3012 | 3825 |
| 1992 | 12996 | 17592 | 23814 | 4226 | 5099 | 6153 | 0.586 | 0.734 | 0.92 | 8384 | 9979 | 11876 | 2677 | 3392 | 4298 |
| 1993 | 12063 | 15873 | 20886 | 3538 | 4278 | 5174 | 0.519 | 0.656 | 0.83 | 7532 | 8980 | 10705 | 2198 | 2778 | 3512 |
| 1994 | 11477 | 15159 | 20024 | 3682 | 4508 | 5519 | 0.506 | 0.638 | 0.805 | 7411 | 8875 | 10629 | 2271 | 2873 | 3634 |
| 1995 | 13516 | 17792 | 23419 | 3200 | 3927 | 4820 | 0.457 | 0.581 | 0.738 | 6672 | 7988 | 9564 | 1886 | 2384 | 3013 |
| 1996 | 16486 | 21761 | 28724 | 3430 | 4237 | 5234 | 0.412 | 0.526 | 0.673 | 7224 | 8676 | 10419 | 1826 | 2288 | 2865 |
| 1997 | 17177 | 22635 | 29828 | 3597 | 4435 | 5469 | 0.409 | 0.52 | 0.661 | 7727 | 9291 | 11172 | 1945 | 2441 | 3064 |
| 1998 | 15099 | 19868 | 26142 | 3859 | 4790 | 5946 | 0.393 | 0.508 | 0.655 | 7989 | 9644 | 11643 | 2063 | 2598 | 3272 |
| 1999 | 14368 | 19050 | 25258 | 4435 | 5565 | 6982 | 0.303 | 0.398 | 0.525 | 8911 | 10834 | 13173 | 1968 | 2479 | 3122 |
| 2000 | 17726 | 23845 | 32076 | 4852 | 6151 | 7796 | 0.243 | 0.331 | 0.45 | 9145 | 11215 | 13753 | 1761 | 2240 | 2850 |
| 2001 | 18163 | 24174 | 32175 | 5808 | 7467 | 9602 | 0.209 | 0.287 | 0.395 | 10468 | 12955 | 16032 | 1766 | 2225 | 2803 |


| Year | Recruitment <br> (thousands) |  |  | SSB (t) |  |  | Fbar (3-6) |  |  | TSB (t) |  |  | Dead catch ( t ) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High | Low | Mid | High |
| 2002 | 18432 | 24688 | 33067 | 6766 | 8806 | 11461 | 0.177 | 0.245 | 0.338 | 11613 | 14565 | 18267 | 1768 | 2218 | 2783 |
| 2003 | 16386 | 22155 | 29955 | 7980 | 10527 | 13886 | 0.143 | 0.202 | 0.287 | 13033 | 16606 | 21158 | 1729 | 2195 | 2787 |
| 2004 | 15564 | 20893 | 28048 | 7999 | 10579 | 13990 | 0.107 | 0.154 | 0.222 | 12678 | 16237 | 20796 | 1339 | 1719 | 2208 |
| 2005 | 13532 | 18137 | 24308 | 7837 | 10324 | 13602 | 0.139 | 0.196 | 0.278 | 12535 | 15893 | 20152 | 1582 | 2007 | 2547 |
| 2006 | 16854 | 22440 | 29878 | 7019 | 9283 | 12276 | 0.151 | 0.212 | 0.296 | 12330 | 15548 | 19605 | 1500 | 1893 | 2388 |
| 2007 | 19368 | 26127 | 35246 | 5793 | 7676 | 10171 | 0.164 | 0.229 | 0.319 | 8841 | 11267 | 14359 | 1290 | 1632 | 2066 |
| 2008 | 16000 | 21325 | 28422 | 5839 | 7714 | 10192 | 0.128 | 0.178 | 0.247 | 10167 | 12835 | 16204 | 1118 | 1405 | 1765 |
| 2009 | 13288 | 17967 | 24293 | 6619 | 8830 | 11778 | 0.095 | 0.134 | 0.189 | 9296 | 11993 | 15473 | 885 | 1132 | 1449 |
| 2010 | 17379 | 23232 | 31057 | 6732 | 8828 | 11577 | 0.135 | 0.191 | 0.269 | 10723 | 13547 | 17114 | 1273 | 1642 | 2118 |
| 2011 | 19947 | 26850 | 36143 | 7887 | 10606 | 14262 | 0.09 | 0.126 | 0.178 | 11999 | 15453 | 19901 | 950 | 1199 | 1513 |
| 2012 | 17988 | 24072 | 32215 | 6617 | 8952 | 12112 | 0.091 | 0.127 | 0.178 | 8638 | 11332 | 14866 | 754 | 951 | 1199 |
| 2013 | 17763 | 23785 | 31849 | 8021 | 10870 | 14730 | 0.061 | 0.086 | 0.121 | 11476 | 14950 | 19476 | 649 | 819 | 1034 |
| 2014 | 19734 | 27392 | 38021 | 8126 | 10961 | 14785 | 0.064 | 0.089 | 0.125 | 11533 | 15014 | 19546 | 689 | 870 | 1097 |
| 2015 | 12734 | 17302 | 23508 | 10482 | 14562 | 20230 | 0.043 | 0.061 | 0.086 | 13716 | 18377 | 24621 | 554 | 705 | 898 |
| 2016 | 10637 | 14284 | 19183 | 15157 | 20726 | 28342 | 0.038 | 0.054 | 0.077 | 19374 | 25733 | 34179 | 737 | 934 | 1184 |
| 2017 | 7741 | 10582 | 14464 | 11551 | 15705 | 21352 | 0.047 | 0.067 | 0.094 | 14340 | 19013 | 25210 | 685 | 866 | 1095 |
| 2018 | 9162 | 12406 | 16799 | 12609 | 17324 | 23801 | 0.048 | 0.068 | 0.097 | 14197 | 19209 | 25990 | 679 | 861 | 1092 |
| 2019 | 8214 | 11245 | 15395 | 11862 | 16532 | 23040 | 0.052 | 0.073 | 0.104 | 13631 | 18600 | 25381 | 627 | 800 | 1020 |
| 2020 | 6011 | 8749 | 12733 | 9617 | 13383 | 18625 | 0.035 | 0.051 | 0.075 | 11360 | 15430 | 20959 | 372 | 481 | 620 |
| 2021 | 6256 | 9261 | 13709 | 9167 | 13064 | 18618 | 0.046 | 0.066 | 0.095 | 10842 | 15012 | 20787 | 449 | 585 | 761 |

Table 24.14 Short-term forecast. Annual catch options. Intermediate year assumptions.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| $F_{\text {ages 3-6 (2022) }}$ | 0.064 | $\mathrm{~F}_{\text {sq }}=\mathrm{F}_{\text {average (2019-2021) }}$ |
| SSB (2023) | 13514 | Tonnes; Fishing at status quo ( $\mathrm{F}_{\text {sq }}$ ). |
| Rage 1 (2022 and 2023) | 11245 | Median resampled recruitment (2015-2021) as esti- <br> mated by a stochastic projection; in thousands. |
| Total catch (2022) | 680 | Tonnes; Fishing at $F_{\text {sq }}$ plus surviving discards. |
| Projected landings (2022) | 322 | Tonnes; Assuming average discard pattern (2019-2021). |
| Projected discards (2022) | $40 \%$ | Tonnes; Assuming average discard pattern (2019-2021). |
| Discard survival rate | 143 | Tonnes; Assuming average discard pattern (2019-2021) <br> where 40\% of the discards survive. |
| Projected surviving discards (2022) | 214 | Tonnes; Assuming average discard pattern (2019-2021) <br> where 40\% of the discards survive. |
| Projected dead discards (2022) |  |  |

Table 24.15. Short-term forecast. Annual catch options. All weights are in tonnes.

| Basis | Total catch <br> (2023) | Projected landings (2023) | Projected Surviving discards (2023) | Projected dead discards (2023) | Total projected discards* (2023) | $\begin{gathered} F_{\text {total }} \\ (2023) \end{gathered}$ | $\mathrm{F}_{\text {projected }}$ landings (2023) | $F_{\text {projected }}$ discards (2023) | $\begin{gathered} \text { SSB } \\ (2024) \end{gathered}$ | \% SSB change *** | \% advice change^ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |  |  |  |
| MSY approach $\mathrm{F}_{\mathrm{MSY}}$ | 2039 | 967 | 429 | 643 | 1072 | 0.196 | 0.061 | 0.135 | 12629 | -6.5 | -30 |
| Other scenarios |  |  |  |  |  |  |  |  |  |  |  |
| $\mathrm{F}_{\text {MSY lower }}$ | 1418 | 672 | 298 | 447 | 745 | 0.133 | 0.042 | 0.091 | 13169 | -2.5 | -52 |
| FMSY upper | 2945 | 1397 | 620 | 929 | 1549 | 0.29 | 0.091 | 0.20 | 11901 | -11.9 | 0.70 |
| $\mathrm{F}_{\mathrm{pa}}$ | 3903 | 1851 | 821 | 1231 | 2052 | 0.40 | 0.126 | 0.28 | 11089 | -17.9 | 33 |
| $\mathrm{F}=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14305 | 5.9 | -100.0 |
| $\mathrm{F}=\mathrm{F}_{\text {lim }}$ | 4649 | 2205 | 978 | 1467 | 2445 | 0.50 | 0.155 | 0.34 | 10468 | -23 | 59 |
| $\mathrm{SSB}_{2024}=\mathrm{B}_{\text {lim }}$ | 13047 | 6187 | 2744 | 4116 | 6860 | 2.33 | 0.73 | 1.60 | 3958 | -71 | 350 |
| $\mathrm{SSB}_{2024}=\mathrm{B}_{\text {pa }}$ | 11370 | 5391 | 2391 | 3587 | 5978 | 1.75 | 0.55 | 1.21 | 5294 | -61 | 290 |
| $\begin{aligned} & \text { SSB }_{2024}= \\ & \text { MSY }_{\text {Brrigger }} \\ & \hline \end{aligned}$ | 6895 | 3269 | 1450 | 2175 | 3625 | 0.81 | 0.25 | 0.56 | 8757 | -35 | 140 |
| Rollover advice | 2925 | 1387 | 615 | 923 | 1538 | 0.29 | 0.091 | 0.20 | 11913 | -11.8 | 0.0 |
| $\mathrm{F}=\mathrm{F}_{2022}$ | 696 | 330 | 146 | 220 | 366 | 0.064 | 0.020 | 0.044 | 13745 | 1.70 | -76 |
| $\mathrm{SSB}_{2024}=\mathrm{SSB}_{2023}$ | 990 | 469 | 208 | 312 | 520 | 0.091 | 0.029 | 0.063 | 13514 | 0 | -65 |

* Dead + surviving projected discards.
** Fprojected discards concerns dead projected discards only.
*** SSB 2024 relative to SSB 2023.
$\wedge$ Advice value for 2023 relative to the advice value for 2022 (2925 tonnes).


Figure 24.1. Irish Sea plaice: Effort and LPUE for commercial fleets from UK (E\&W), Ireland and Belgium.


Figure 24.2a. Landings-at-age data (left) and mean standardised proportion-at-age (right, black bubbles are positive values and orange bubbles are negative). Mean standardised proportion-at-age $=[$ (proportion-at-age in year) - mean (proportion-at-age over all years) ] / STDEV(proportion-at-age over all years).


Figure 24.2b. Discards-at-age data (left) and mean standardised proportion-at-age (right, black bubbles are positive values and orange bubbles are negative). Mean standardised proportion-at-age = [ (proportion-at-age in year) - mean (pro-portion-at-age over all years) ] / STDEV(proportion-at-age over all years).


Figure 24.3. Left: UK(E\&W)-BTS-Q3 (extended area) cpue by age (circles in 2021 due to missing data in 2020). Right: standardised indices of SSB derived from NIGFS-WIBTS, biomass from UK(E\&W)-BTS-Q3 (extended area) (black circle in 2021 due to missing data in 2020) and the SSB estimates from the Annual Egg Production Methods (circles, right).


Figure 24.4. Make up of catch estimates from InterCatch.


Figure 24.5. Catch sampling for landings (left) and discards (right) by country and gear type. Gears contributing less than 1 tonne are excluded for clarity.


Figure 24.6. Northern Irish Groundfish Survey SSB indices split into spring (left hand panels) and autumn (right hand panels) sampling by western strata (1-3), eastern strata (4-7) and total survey area (strata 1-7) with confidence intervals ( $\pm 1$ standard error, vertical lines).

Biomass Eastern Irish Sea


Biomass Western Irish Sea


Figure 24.7. Trends in biomass indices (kg per km towed) the NIGFS-WIBTS-Q1 and -Q4 (blue and red lines respectively) in the eastern Irish Sea (top) and the western and southern Irish Sea (bottom). Also shown (green dots, right axis) are the estimates of SSB from the Annual Egg Production Method (AEPM) from Armstrong et al. (2001).


Figure 24.8. Selectivity of the fishery split into the landed (dashed) and discarded (solid) components as estimated by the SAM model, where the $x$-axis shows age and the $y$-axis gives the fishing mortality-at-age split by the proportion of fish (by number) discarded and landed at-age.


Figure 24.9. Catchability for the UK (E\&W)-BTS-Q3 extended index by age, NIGFS-WIBTS-Q1 and NIGFS-WIBTS-Q4 as estimated by the SAM model.


Figure 24.10. Residuals in fits to catch and survey data from the baseline model. Expected values were estimated by the SAM model.


Figure 24.11. Comparison of the standardised age 1 index from the UK (E\&W)-BTS-Q3 extended area (purple and purple cycle due to missing data in 2020) and the standardised recruitment (green dashed line) estimated by the SAM model.

Biomass estimates


Figure 24.12. SAM model estimates of mean standardised SSB (orange line) overlaid with standardised NIGFS in spring (red) and autumn (blue dashed) relative SSB indices, standardised biomass (ages 1-4) from the UK(E\&W)-BTS (black solid line and black cycle due to missing data in 2020) and AEPM SSB index (circles, right axis). Standardized: minus mean and divided by standard deviation.


Figure 24.13. Modelled SSB (tonnes, top left), recruitment (thousands, bottom left), F bar $^{\text {(ages 3-6, bottom right) catch }}$ tonnage (bottom right) using the SAM model. Error dashed lines indicate $\mathbf{2 \times}$ standard deviation.


Figure 24.14. Retrospective assessments for years 2011-2021 from the baseline model. SSB (tonnes, top left), recruitment (thousands, bottom left), $\mathrm{F}_{\text {bar }}$ (ages 3-6, bottom right) catch tonnage (bottom right). Error dashed lines indicate $\mathbf{2 \times s t a n d - ~}$ ard deviation.


Figure 24.15. Retrospective assessments for years 2011-2021 from the baseline model, showing final 11 years. SSB (tonnes, top left), recruitment (thousands, bottom left), $\mathrm{F}_{\text {bar }}$ (ages 3-6, bottom right) catch tonnage (bottom right). Error dashed lines indicate $2 \times$ standard deviation.

## 25 Plaice (Pleuronectes platessa) in Division 7.e (western English Channel

## Type of assessment in 2021

Last year's assessment report is available at: https://www.ices.dk/sites/pub/Publication\ Re-ports/Expert\ Group\ Report/Fisheries\ Resources\ Steer-
ing\%20Group/2021/WGCSE publication\%20with\%20multi-
ple\%20files/WGCSE2021 25\%20Plaice\%207e.pdf.

## ICES advice applicable to 2022

Last year's advice is available at https://doi.org/10.17895/ices.advice. 7822 and stated:
ICES advises that when the precautionary approach is applied, catches in 2022 should be no more than 1742 tonnes.

### 25.1 Impact of the COVID-19 pandemic

The plaice in Division 7.e stock, its fishery, and data sampling were largely unaffected by the implications of the COVID-19 pandemic in 2021.

### 25.2 ICES Transparent Assessment Framework

The Division 7.e plaice stock is included in the ICES Transparent Assessment Framework (TAF, https://taf.ices.dk, https://github.com/ices-taf). All WGCSE assessments since 2019 are available from the ICES TAF GitHub page (please note, access to these repositories is so far restricted to ICES and members of WGCSE). The current WGCSE 2022 assessment is available from https://github.com/ices-taf/2022 ple.27.7e assessment.
All changes since last year's assessment can be accessed with the following link: https://github.com/ices-taf/2022 ple.27.7e assessment/compare/de247c6...main.

The TAF repository includes all input data, R scripts for processing data, preparing and running the stock assessment and advice rule, and scripts for creating all figures and tables presented in this report.

### 25.3 General

### 25.3.1 Stock description and management units

The ICES advice for this stock is strictly that for ICES Division 7.e, called the western English Channel. The TAC area does not correspond to the stock area, given that it includes the larger component of 7.d (eastern English Channel). However, WKFLAT 2010 (ICES, 2010) found that a significant proportion of the catches of the 7.e stock are taken in the adjacent division during the spawning period. Plaice is not the main target species in 7.e, and it is generally taken as bycatch in fisheries targeting sole.


TAC area $=7 . \mathrm{d}-\mathrm{e}$; Assessment area $=7 . \mathrm{e}$.

### 25.3.2 Management applicable to 2021 and 2022

There are technical measures in operation, including a minimum 80 mm mesh size and a minimum landings size $(27 \mathrm{~cm})$ for this species.

The TAC and the national quotas by country for 2021 (in tonnes)

| Species | Plaice <br> Pleuronectes platessa | Zone: <br> 7d and 7e <br> (PLE/7DE.) |
| :--- | :--- | :--- |
| Belgium | 1537 | Analytical TAC |
| France | 6850 | Article 8(2) of this Regulation applies |
| Union | 8387 |  |
| United Kingdom | 3533 |  |
| TAC | 11920 |  |

(Source: Council Regulation (EU) 2021/1239, EU, 2021).

The TAC and the national quotas by country for 2021

| Species | Plaice <br> Pleuronectes platessa | Zone: <br> 7d and 7e <br> (PLE/7DE.) |
| :--- | :--- | :--- |
| Belgium | 1310 | Analytical TAC |
| France | 4366 | Article 8(2) of this Regulation applies |
| Union | 5676 |  |
| United Kingdom | 2717 |  |
| TAC | 9138 |  |

(Source: Council Regulation (EU) 2022/515, EU, 2022).

### 25.3.3 Landing obligation

The EU landing obligation was phased in between 2019 and 2021 for plaice in 7.e with a discard plan defined in the Commission Delegated Regulation (EU) 2018/2034 (EU, 2018) and Commission Delegated Regulation (EU) 2019/2239 (EU, 2019) and referring to Regulation (EU) No 1380/2013 (EU, 2013). According to this discard plan, the landing obligation applies to plaice in 7.e. since 1 January 2019. There are, however, survivability exemptions for plaice when caught with specific gears. This includes all (a) trammelnets (gear codes GTR, GTN, GEN, GN) and (b) otter trawls (gear codes OTT, OTB, TBS, TBN, TB, PTB, OT, PT, TX). Furthermore, Commission Delegated Regulation (EU) 2018/2034 (EU, 2018a) set a provisional exemption for 2019, including BT2 beam trawls (i.e. 80 mm to 120 mm mesh size) for (c) vessels with a maximum engine power greater than 221 kW and fitted with a flip-up rope or benthic release panel, and (d) for vessels with a maximum engine power of 221 kW or a maximum length of 24 m , when fishing within 12 nautical miles of the coast and with average tow durations of no more than 1:30 hours (Commission Delegated Regulation (EU) 2018/2034, Article 6, EU, 2018a).

This provisional exemption was extended to 2020 in Commission Delegated Regulation (EU) 2019/2239 (EU, 2019) and points (c) and (d) extended to beam trawls irrespective of mesh size (Commission Delegated Regulation (EU) 2019/2239, Article 6, EU, 2019).

Prior to introducing the landing obligation, a substantial part of the plaice 7.e catches has been discarded and not accounted for in the stock assessment. In the first year of the phasing in of the landing obligation, the exemptions are likely to cover most of the plaice catches, and the impact on fishing or stock assessment is likely to be negligible. In the following years of the discard plan, the situation should be closely monitored because of potential changes in the landings data and composition, which might affect the stock assessment.

### 25.4 Data

### 25.4.1 InterCatch

International catch data are collated on the ICES InterCatch platform (https://intercatch.ices.dk). In the Western English Channel, plaice is taken mainly as bycatch in beam trawls targeting sole and anglerfish. In 2021, $71.2 \%$ of the landings were taken by beam trawls, $21.7 \%$ by otter trawls, $4.1 \%$ by gillnets and $2.9 \%$ by other gears. Of the total international landings, $85.6 \%$ were taken by the UK, $6.6 \%$ by France, $7.7 \%$ by Belgium, and $0.1 \%$ by the Netherlands (Table 25.1, Figures 25.1 and 25.2).

This stock is the smaller of the two plaice stocks that make up the larger TAC Area 7.d-e. The official landings from this stock amounted to $15 \%$ of the TAC in $2019,14 \%$ in 2020 and $15 \%$ in 2021. The combined catches of plaice in 7.d-e accounted for $50 \%$ of the TAC in $2019,74 \%$ in 2020, and $34 \%$ in 2021.

### 25.4.2 Landings

National landings data reported to ICES and estimates of total landings used by the Working Group are given in Table 25.1. Total international plaice landings in Division 7.e were 1275 t in 2020 and 1331 t in 2021, an increase of $4 \%$.

In addition to the estimated 2021 landings for Division 7.e, an extra of 72 tonnes ( 98 tonnes in 2020) were added from the 7. d plaice stock representing an adjustment for migration of $15 \%$ of the mature component of quarter 1 landings between the two divisions. This process was agreed at WKFLAT 2010 (ICES, 2010), and the migration correction was revised at WKPLE 2015 (ICES, 2015a). The process is described in the Stock Annex. A reciprocal correction is made to the 7.d plaice stock. Figure 25.3 shows the total annual landings split by divisions 7.e and 7.d.

### 25.4.3 Discards

Although discards have not been used in the assessment of 7.e plaice historically, some discard data are available. Discard tonnages are available within InterCatch and were provided by the UK(E\&W) for the years 2012-2021, France for 2014-2021, Belgium for 2012-2013 and 2015-2021, the Netherlands for 2015-2019 and Ireland for 2017-2021 (zero discards reported).

Discard coverage and sampling are generally at a high level for this stock. Discard estimates were provided for $88 \%$ of the landings in $2019,23 \%$ in 2022 (impacted by Covid restrictions on sampling), and $91 \%$ in 2021. Of these discard estimates, age samples were provided for $89 \%$ in each of the years 2019, 2020, and 2021.

In analogy to the landings, the discards are also uplifted by a migration correction from 7.d. For 2021, 95 tonnes ( $15 \%$ of the mature quarter 1 plaice discards in $7 . d$ ) were added, resulting in total discards of 211 tonnes for the 7.e plaice stock.

For historical consistency reasons, Figure 25.4 shows various discard rates for plaice in $7 . e$. Since WGCSE 2017, the discard rate has been calculated as the contribution of total discards (raised, including migration correction) to the total plaice landings (including migration component), and this discard rate was $13.1 \%$ in 2021 (Figure 25.5).

### 25.4.4 Sampling

Sampling levels for this stock have been high in recent years.
This year, all nations (apart from Scotland and the Belgian beam trawl fleet) provided data disaggregated by fleet and by quarter, and these were all uploaded into the ICES InterCatch database. Quarterly age compositions for landings in 2021 were available from the UK (England) and were provided for four fleets (GNS_DEF_all_0_0_all Q3, Q4; OTB_DEF_>=120_0_0_all Q1, Q2, Q3 Q4; OTB_DEF_70-99_0_0_all Q1, Q2, Q3, Q4; and TBB_DEF_70-99_0_0_all Q1, Q2, Q3, Q4), Belgium for one fleet (TBB_DEF_70-99_0_0_all, annual), and France for two fleets (OTB_DEF_100-119_0_0 Q3; OTB_DEF_70-99_0_0 Q3). Sampling levels for landings and discards were good for plaice and age samples covered $91 \%$ of the total reported international landings and $89 \%$ for discards. Figure 25.6 visualises age samples and gives details about the number of samples and age readings in 2021.

Additional landings data were available by quarter/fleet from Belgium, France, Ireland (0 landings), the Netherlands ( $<1 \mathrm{t}$ ), UK (E+W, Guernsey) and UK Scotland (annual, $<1 \mathrm{t}$ ). These datasets were aggregated to an international age-structured catch using the ICES InterCatch platform.

Length compositions were provided by the UK (E\&W), Belgium and France and covered 91\% of all submitted landings. Figure 25.7 visualises age samples and gives details about the number of samples and length readings in 2021.

An additional age composition representing the migration adjustment ( $15 \%$ of the mature component of quarter 1 catches for $7 . \mathrm{d}$ ) was supplied on request by the WGNSSK stock coordinator for the 7.d plaice stock.

The method for deriving the international catch numbers and the calculation of the catch and stock weights-at-age has been fully described in the Stock Annex, Section B1. Landings numbers-at-age (including the migration element) are given in Figure 25.9. Landings and stock weights-at-age are plotted in Figures 25.9 and 25.10.

Catch weights are assumed to be mid-year values, and stock weights are interpolated back (in year) to 1 January, as standard for this stock (Figure 25.9).

### 25.4.5 Revisions

No revisions to data prior to 2021 were provided in 2022.

### 25.4.6 Biological

The natural mortality and the maturity ogives used were identical to previous assessments and as described in the Stock Annex.

### 25.4.7 Surveys

Two surveys currently provide abundance estimates to the Working Group (Figure 25.11, Figure 25.12, Table 25.2, and Figure 25.13 for internal consistency).

### 25.4.7.1 UK Fisheries Science Partnership

The UK Western Channel sole and plaice survey (previously called Fisheries Science Partnership survey; UK-FSP, quarter 3, ICES survey code B4381, Burt et al., 2022) conducted another survey of sole and plaice abundance in the Western English Channel in 2021. The survey uses two 4 m beam trawls with 80 mm nominal codend mesh and focuses on the area around the English coast. 83 out of 90 tows were completed in 2021. 497 plaice otoliths were collected for aging in 2021.

For this survey, catch rates are reported standardised as numbers and biomass (kg) per hour per meter of beam length.

The plaice biomass from the FSP survey has been decreasing since 2014 and 2021 is the first increase in biomass since then. Plaice were encountered at $95 \%$ of the survey stations, with a similar distribution to previous years, with greater numbers in Bigbury Bay and Lyme Bay (Figure 25.14, Burt et al., 2022). Plaice catches comprised mainly fish aged 2-4.

Internal consistency (cohort tracking) is acceptable for this survey for ages three and above (Figure 25.13).

### 25.4.7.2 Q1SWBeam

The second survey used for sole is the Quarter 1 South West Beam trawl (Q1SWBeam, also called Q1SWECOS, ICES survey code B2732), which started in 2006. This survey deploys two 4 m beam trawls and uses a fully random stratified approach. In contrast to the FSP survey, the Q1SWBeam covers the entire western English Channel and, if conditions permit, adjacent areas. In 2021, 78 out of 81 tows were completed in the western English Channel.

For this survey, only standardised abundance (numbers-at-age) are reported regularly.
Plaice are mainly caught along the English coast with small or zero catches elsewhere in the western English Channel. Plaice catches are mainly comprised of fish aged 3 and 4. Cohort tracking is moderate for ages 3 to 7 (Figure 25.13). There is no statistically significant correlation ( $\mathrm{p}<0.05$ ) between the catches-at-age 2 and 3 , between 6 and 7,7 and 8 , and 8 and 9 .

### 25.4.7.3 Commercial fleet effort and LPUE

UK(E\&W) beam trawl and otter trawl time-series are shown in Figure 26.15.
UK(E\&W) beam trawl effort is relatively stable at high levels since the early 2000s but the landings increased substantially between 2015 and 2016 but have been decreasing since.

UK(E\&W) otter trawl effort (days fished-GRT corrected) has declined since 1989 to very low levels in recent years. In 2016, this fleet reported 0 effort and no landings, i.e. there is no LPUE value for 2016. The reason for this is that the LPUE otter trawl index is calculated only with vessels of at least 12 m length, and in 2016 only smaller vessels deploying otter trawls reported any activity. Due to a change in the database system, no values are available for the otter trawl fleet after 2016, consistent with the earlier period.

### 25.5 ICES advice considerations

Since 2015, this stock has been classified as category 3 data-limited by ICES for advice purposes. Until 2021, ICES advice was provided following the 2012 data-limited stocks framework (ICES, 2012) and the catch advice was based on the " 2 over 3 " rule (method 3.2 in ICES, 2012). The stock trend used in the 2 over 3 rule was based on the SSB estimate of a landings only XSA assessment.

In 2022, ICES rolled out new data-limited advice rules for category 3 stocks developed by ICES WKLIFE X (ICES, 2020). Consequently, the advice for this plaice stock is based on the new framework (ICES, 2022).

### 25.6 Choice of method for providing advice

The choice of method for providing advice for plaice is based on the new ICES technical guidelines for stocks in categories 2 and 3 (ICES, 2022).

The first step in choosing a method is considering whether a surplus production model (such as SPiCT, Pedersen and Berg, 2017) has been fit successfully to the stocks and whether the model fit meets acceptance criteria. In previous years, WGCSE has conducted extensive explorations of SPiCT for this plaice stock. However, this was not successful. The model exhibited very large uncertainty and was sensitive to the first and last year of data used in the model. Alternative model configurations, fixing parameters, or including priors on model parameters did not lead to a suitable model. Consequently, SPiCT was rejected by WGCSE as this model does not appear to be able to model the stock dynamics appropriately.

For this stock, an index of abundance, as well as length data, are available. Following ICES (2022), the choice of the method then depends on the individual growth rate (quantified by the von Bertalanffy growth model parameter $k$ ).

### 25.6.1 von Bertalanffy growth parameters

Individual growth parameters were estimated by fitting a von Bertalanffy growth model to stock-specific age-length data:

$$
L_{t}=L_{\infty}\left(1-e^{-k\left(t-t_{0}\right)}\right)
$$

For this stock, age-length data were available from UK commercial catch sampling and from the Q1SWBeam survey. The commercial data include older fish but are lacking younger ages. Conversely, the Q1SWBeam data include younger ages but lack older individuals. Consequently, the two data sources were combined. This approach is appropriate because the fish come from the same area (ICES Division 7.e) and were caught with similar gears (the Q1SWBeam uses a beam trawl with 80 mm mesh size, the majority of the commercial fishery is taken by beam trawls with mesh size 70-99 mm). The growth model was fit to annual data for 2006-2021 (Figure 25.16, Table 25.3). There has been little variability in the growth model in recent years. For more robust model estimates, the von Bertalanffy model was also fit to the combined data from the last five years (2017-2021), where the data from different years and data sources were weighted equally. This resulted in the following von Bertalanffy individual growth parameters:

| von Bertalanffy parameter | $L_{\infty}$ | $k$ | $t_{0}$ |
| :--- | :---: | :---: | :---: |
| value | 58.5 cm | 0.11 year $^{-1}$ | -2.3 years |

Following ICES (2022), this means that the "rfb rule" should be applied because $k<0.32$ year $^{-1}$.

### 25.7 Application of the rfb rule

In 2022, WGCSE applied the rfb rule (Fischer et al., 2020, 2021a,b, ICES, 2020, 2022) to plaice in Division 7.e.

The rfb rule is an empirical harvest control rule meant to be applied to empirical data and not estimates from a population model. Consequently, the previously used XSA assessment was discontinued and the rfb rule was directly applied to empirical data (survey and catch data).

The rfb rule is defined as

$$
A_{y+1}=A_{y} r f b m
$$

where $A_{y+1}$ is the new advised catch for year $y+1, A_{y}$ the previously advised catch, $r$ a biomass ratio following the trend in a biomass index, $f$ a fishing pressure proxy using catch length data, $b$ a biomass safeguard protecting against low stock size, and $m$ a multiplier ensuring long-term precautionary exploitation.

The rfb rule includes a conditional and asymmetric uncertainty cap restricting changes in $A_{y+1}$ relative to $A_{y}$ to $+20 \%$ and $-30 \%$, but is only implemented if $b=1$. Furthermore, the rfb rule provides biennial catch advice, i.e. once set, the advice is kept for two years.

The rfb rule was applied according to the 2022 ICES guidelines (ICES, 2022) and the following sections describe how the components of the rfb rule were derived.

### 25.7.1 Component $\boldsymbol{A}_{\boldsymbol{y}}$

The rfb rule derives the catch advice by adjusting the previously advised catch. However, the rfb rule is meant to adjust realised catches affecting a fish stock. Figure 25.17 shows a comparison of historical catch and ICES advised catches. In recent years, the advised catch has been above the realised catches, but the advised catches have been decreasing. The estimated catch for 2021 is close to the advised catch for 2022. Therefore, using the ICES catch advice for 2022 in the rfb rule appears appropriate:

$$
A_{y}=A_{2021}=1742 \text { tonnes } \quad \text { (source: } \underline{\text { ICES advice 2021; ICES, 2021) }}
$$

### 25.7.2 Component $r$

Component $r$ of the rfb rule informs on the biomass trend of the stock. Two survey indices exist for plaice: UK-FSP and Q1SWBeam. The UK-FSP survey was selected for application in the rfb rule because (1) it occurs later in the year in quarter 3 and therefore provides a more recent estimate of stock biomass, (2) it had a higher contribution compared Q1SWBeam in previous stock assessment model fits and shows better internal consistency (cohort tracking, see Figure 25.13) and appears more robust and less susceptible to noise in the data, (3) it covers the main habitat for plaice, and (4) it is the only survey for which a time-series of standardised biomass estimates (in kg per hour per metre of beam) are available. For the biomass index, only ages 2-8 were considered because these were the ages previously selected in age-structured assessments and younger as well as older fish might not be fully selected and likely provide more noise than signal.
Component r of the rfb rule is calculated as (ICES, 2022):

$$
r=\frac{\sum_{i=y-2}^{y-1} I_{i} / 2}{\sum_{i=y-5}^{y-3} I_{i} / 3}=\frac{\sum_{i=2020}^{2021} I_{i} / 2}{\sum_{i=2017}^{2019} I_{i} / 3}=\frac{0.81}{1.10}=0.74
$$

where $I$ is the biomass of the UK-FSP survey, aggregated over ages 2-8. Figure 25.18 illustrates the biomass index and the calculation of $r$.

### 25.7.3 Component $f$

Component $f$ of the rfb rule is a proxy for the fishing pressure and uses catch length data. Catch length distributions were generated from InterCatch following the same raising and allocation principles as for the age distributions. Figure 25.19 illustrates international catch (landings and discards) length distributions for 2014-2021.

Component $f$ requires the definition of length at first capture $L_{c} . L_{c}$ was calculated annually for 2014-2021 following the ICES guidelines (ICES, 2022), which define $L_{c}$ as the first (smallest) length class for which the catch numbers are at or above the mode of the distribution (the length class with the highest catch numbers). $L_{c}$ showed little variability over the years ( $26-27 \mathrm{~cm}$, Table 25.4, Figure 25.19) and the average over 2017-2021 was used as the final $L_{c}=26.4 \mathrm{~cm}$.

The annual mean length in catch $L_{\text {mean }}$, for length classes above $L_{c}$ was calculated as the mean of length classes above $L_{c}$, weighted by the catch numbers per length class (Table 25.4, Figure 25.19).

The mean catch length can be compared to an MSY proxy reference length to infer fishing pressure on the stock. $L_{F=M}$ is used as a reference length (ICES, 2022), and calculated as:

$$
L_{F=M}=0.75 L_{c}+0.25 L_{\infty}=0.75 \times 26.4 \mathrm{~cm}+0.25 \times 58.5 \mathrm{~cm}=34.4 \mathrm{~cm}
$$

This is an approximation and assumes $M / k=1.5$, and was used for plaice because natural mortality $M$ is not known reliably. $L_{\infty}$ and $L_{c}$ used here are average over the last five years. This ensures a more robust estimation and means that $L_{F=M}$ is assumed constant over time.
$L_{\text {mean }}$ can then be compared to $L_{F=M}$ (Figure 25.19, 25.20). $L_{\text {mean }}$ was below $L_{F=M}$ for 2014-2021, which indicates overfishing of the plaice $7 . e$ stock, a perception that is supported by exploratory age-structured stock assessment models.

Component $f$ of the rfb rule is calculated as:

$$
f=\frac{L_{\text {mean }}}{L_{F=M}}=\frac{L_{2021}}{L_{F=M}}=\frac{32.3 \mathrm{~cm}}{34.4 \mathrm{~cm}}=0.94
$$

### 25.7.4 Component $b$

The biomass safeguard (component $b$ ) protects against low stock size and includes a biomass index trigger value ( $I_{\text {trigger }}$ ). The same biomass index as for component $r$ is used for $b$. In the absence of better knowledge, $I_{\text {trigger }}$ is based on the lowest observed biomass index value ( $I_{\text {loss }}$ ):

$$
I_{\text {trigger }}=I_{\text {loss }} \times 1.4=I_{2007} \times 1.4=0.28 \mathrm{~kg} \mathrm{hr}^{-1} \mathrm{~m} \mathrm{beam}^{-1} \times 1.4=0.39 \mathrm{~kg} \mathrm{hr}^{-1} \mathrm{~m} \mathrm{beam}^{-1}
$$

Component $b$ is then calculated as (ICES, 2022):

$$
b=\min \left\{1, \frac{I_{y-1}}{I_{\text {trigger }}}\right\}=\left\{1, \frac{I_{2021}}{I_{\text {trigger }}}\right\}=\min \left\{1, \frac{1.03}{0.39}\right\}=1
$$

### 25.7.5 Component $\boldsymbol{m}$

The multiplier $m$ is set depending on the von Bertalanffy parameter $k$ (ICES, 2022):

$$
m= \begin{cases}0.95, & \text { if } k<0.20 \text { year }^{-1} \\ 0.90, & \text { if } 0.20 \leq k<0.32 \text { year }^{-1}\end{cases}
$$

and because for plaice $k<0.20$ year $^{-1}, m$ is set to:

$$
m=0.95
$$

### 25.7.6 Combining the rfb rule's components

The catch advice with the rfb rule is calculated as

$$
A_{y+1}=A_{y} r f b m=1742 \text { tonnes } \times 0.74 \times 0.94 \times 1 \times 0.95=1146 \text { tonnes }
$$

However, the rfb rule is used in combination with a conditional uncertainty cap ( $+20 \%,-30 \%$ ), implemented only when $b=1$. For plaice in $2022, b=1$, and the uncertainty cap was considered. The change in the advised catch would be a reduction of more than $30 \%(r \times f \times b \times m=$ $0.67 \times 0.94 \times 1 \times 0.95=0.66$, i.e. $-34 \%)$. This means the final catch advice is capped:

$$
A_{y+1}=A_{y} \times 0.7=1742 \text { tonnes } \times 0.7=1219 \text { tonnes }
$$

This leads to the final catch advice for the 7.e plaice stock of 1219 tonnes in each of the years 2023 and 2024 (biennial advice).

Table 25.5 summarises the ICES catch advice and also provides the advice corresponding to ICES Division 7.e.

### 25.8 Legacy XSA assessment

In previous years, WGCSE conducted a landings-only XSA assessment and the SSB estimates from this assessment were used to inform the 2 over 3 rule. This assessment is not used anymore for the ICES advice because the rfb rule is applied to empirical data. Furthermore, the XSA assessment does not include discards despite substantial discarding for 7.e plaice. Nevertheless, this "legacy" XSA assessment has been updated at WGCSE 2022. The results are shown in Figure 25.21, residuals in Figure 25.22, and a retrospective analysis in Figure 25.23.

The output of this XSA assessment should be considered with caution because it ignores discards and exhibits a large retrospective pattern.

### 25.9 Exploratory SAM assessment

Since last year (WGCSE 2021), an exploratory assessment using the state-space stock assessment model SAM (Nielsen and Berg, 2014) has been conducted in parallel. This assessment includes discards (including reconstructed discards prior to 2012 where no InterCatch discard estimates are available). This assessment is briefly presented here. Please note that this SAM assessment for 7.e plaice was never benchmarked or reviewed, is not used for providing advice, and should be considered exploratory only.

The model uses total catches (including discards) from 1980-2021, the two survey indices (Q1SWBeam, 2006-2021, ages 2-9; UK-FSP, 2003-2021, ages 2-8) and other input data previously used for XSA (biological data such stock weights, natural mortality, maturity, etc.). The assessment uses SAM's default model configuration; the only exception is that the last two ages for the UK-FSP survey and the last three ages for the Q1SWBeam survey are linked to mimic the model configuration previously adopted for XSA.

Figure 25.24 shows a summary of the results of this SAM assessment. The reference points were derived at WGCSE by running an EqSim model. This SAM assessment indicates strong overfishing ( $\mathrm{F}>\mathrm{F}_{\mathrm{MSY}}$ ).

Figure 25.25 illustrates model diagnostics. The SAM model exhibits a minor retrospective pattern (Figure 25.25a, Mohn'r rho for SSB is $+8.3 \%$, for $\mathrm{F}-9.8 \%$ ), minor rescaling when removing either survey (Figure 25.25b), residuals (Figure 25.25c) and process residuals (Figure 25.25d) appear appropriate, is robust to alternative initial parameters in the model fitting process (Figure 25.25 e ), and simulating data from the model and refitting SAM (Figure 25.25f) leads to replicates within the estimated uncertainty boundaries.

### 25.10 Management plans

There is no management plan in place for this stock apart from the EU multiannual plan for the region.

### 25.11 Uncertainties

A degree of uncertainty exists over the landings statistics for this stock, given that mature plaice migrate between 7.d and 7.e during the spawning period. The current assessment applies a spawning migration correction that reallocates $15 \%$ of quarter 1 landings for the mature proportion of the catch from 7.d to 7.e. Consequently, the assessment results depend on the mixing rate assumption estimated from existing tagging data. Further work is required to examine the stock structure and the mixing rate during the spawning period. Additional data are also needed to determine if the current mixing rate remains valid given the increased abundance of plaice stocks in the English Channel in recent years.

France had to revise their 2018 discard data previously due to inaccuracies in the raising procedure. However, the same procedure has already been used in previous years, which have not been revised.

### 25.12 Recommendations for the next benchmark

The 7.e plaice is relatively data-rich for an ICES category 3 data-limited stock. There are no major concerns for this stock to provide advice following category 3 methods. However, should the stock be upgraded in the future, this section provides an overview of potential issues. An updated issue list is kept on the ICES system for rolling issues (https://sid.ices.dk/Manage/rollingissues.aspx).

Most of the following points have been described in previous reports and are repeated this year.
A benchmark assessment was developed for this stock at WKFLAT 2010 (ICES, 2010), and an inter-benchmark meeting (IBPWCFlat2, ICES, 2015b) subsequently convened in 2015 to revise the input data and update the XSA assessment settings. Nevertheless, any future benchmark meeting will need to consider the following issues.

Since 2017, ICES asked for the additional application of data-limited methods for category 3 stocks. This had massively increased the workload for the stock coordinator and assessor, but with little benefit for this stock. Upgrading this stock to category 1 was considered desirable and feasible within a reasonable time frame. In 2022, due to the application of the rfb rule, the need to apply additional exploratory data-limited methods has disappeared.

The decisive reason for downgrading the stock to category 3 in 2015 was unacceptable retrospective patterns in the XSA assessment. For the application of an analytical assessment, the following issues need to be considered:

- A discard time-series should be developed and included in the assessment as discarding was substantial in recent years.
- Discards, including age compositions, are now routinely estimated within InterCatch and exist for 2012-2021. Some UK discards data prior to 2012 exist but are not used so far. The discard time-series should be extended back in time, as it has been done for other plaice and similar stocks. An exploratory assessment with a historical discard guestimation has been conducted since WGCSE 2020 and indicates considerably higher fishing mortality in recent years than the previous landings only assessment.
- Including discards in the assessment might require a reparameterisation of XSA settings and exploring alternative age-structured assessment models, such as SAM.
- Biological data such as natural mortality and maturity ogives are time-invariant in the current assessment and borrowed from other plaice stocks (divisions 7.fg and 7.a). There have been benchmarks for other plaice stocks, and a similar approach could be made for plaice in 7.e. The natural mortality used for plaice in 7.e was originally borrowed from
plaice in Division 7.a. The values for plaice in 7.a have been changed recently, but the original values are still used for plaice in 7.e.

Furthermore, the following points should be considered:

- Smoothing of stock and catch weights. The raw catch weights are corrected for migration from 7.d and then smoothed using a polynomial function of second degree. Even though the fit seems quite reasonable, different, more appropriate methods should be evaluated.
- Abundance estimates derived from the UK FSP-7e and Q1SWBeam surveys included in the assessment are spatially restricted to the same areas as the commercial tuning fleets, and therefore little population abundance information exists along the French coast. CPUE estimates from additional research surveys in French coastal waters would improve the robustness of future assessment outputs.
- Cohort tracking in the Q1SWBeam index is only mediocre.
- Investigate the addition of age composition information from the French and Belgian fleets. These fleets collectively accounted for about $16 \%$ of the total landings of this stock. In particular, the inclusion of French data would add information on the stock dynamics on the French coast.
- In 2019/2020 there was a revision of the Q1SWBeam survey index, with changes to station validities and the calculation of the index. This should be further reviewed.
- France revised the 2018 discard data due to an issue with the raising procedure. However, the same procedure has been used for previous which have not been revised.
- The landings only XSA assessment is again exhibiting larger retrospective patterns. These patterns are reduced when discards are included in the assessment.


### 25.13 Management considerations

The stock unit (Division 7.e) does not correspond with the management unit (divisions 7.d and 7.e), and this divisional mismatch hampers the effective management of plaice in the western English Channel. However, some provision must be made to consider the effective management of adjacent plaice stocks, given that components of the 7.e stock are also taken during the spawning period in 7.d. WKPLE 2015 revised the established migration correction so that $15 \%$ of quarter 1 landings for the mature proportion of the catch are reallocated from 7.d to 7.e, and the associated age composition is applied to plaice 7.e.

Due to migration patterns, catches of this stock also occur in Division 7.d during the spawning period; therefore, to be consistent with the advised catch for the Division 7.e plaice stock, the actual catches of plaice in Division 7.e should be lower than the advised catch for the stock. ICES has calculated the corresponding actual catches in Division 7.e, assuming that the proportion of Division 7.e stock catches taken in Division 7.d remains as in previous years (i.e. $9 \%$, the average of 2003-2021, taking the age structure of the population into account). As the mixing rate of the two plaice stocks is uncertain, this calculation provides only an approximation.

The total allowable catch (TAC) for the management area has not always been following scientific recommendation for the combined divisions 7.e and 7.d. for 2016 has been doubled compared to 2015 but was reduced slightly in the following years.

### 25.14 Stock assessment audit

The 2021 stock assessment for this stock has been audited internally within WGCSE, and no issues were found.

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### 25.16 Tables

Table 25.1. Plaice in 7.e. History of official landings by country and ICES estimates of landings and discards. All weights are in tonnes.

| $\begin{aligned} & \text { ぁ } \\ & \stackrel{y}{0} \end{aligned}$ | $\begin{aligned} & E \\ & \frac{E}{D} \\ & \frac{0}{D} \\ & \hline D \end{aligned}$ |  | $\begin{aligned} & \underset{\text { U }}{0} \\ & \stackrel{\pi}{4} \end{aligned}$ |  | $\begin{aligned} & \stackrel{\sim}{\omega} \\ & \frac{1}{ \pm} \\ & \hline \end{aligned}$ | $\bar{\pi}$ 0 0 0.0 0.0 4 0 0 | $\begin{aligned} & \stackrel{*}{斤 0} \\ & \stackrel{7}{\circ} \end{aligned}$ |  |  | $\begin{aligned} & \stackrel{*}{*} \\ & \stackrel{n}{n} \\ & \stackrel{n}{0} \\ & \stackrel{H}{n} \\ & \ddot{0} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 5 | - | 323 | 312 | - | 640 | 640 | - | 640 |  |
| 1977 | 3 | - | 336 | 363 | - | 702 | 702 | - | 702 |  |
| 1978 | 3 | - | 314 | 467 | - | 784 | 784 | - | 784 |  |
| 1979 | 2 | - | 458 | 515 | - | 975 | 977 | - | 977 |  |
| 1980 | 23 | - | 325 | 609 | 9 | 966 | 1079 | 99 | 1178 |  |
| 1981 | 27 | - | 537 | 953 | - | 1517 | 1501 | 175 | 1676 |  |
| 1982 | 81 | - | 363 | 1109 | - | 1553 | 1688 | 190 | 1878 |  |
| 1983 | 20 | - | 371 | 1195 | - | 1586 | 1495 | 219 | 1714 |  |
| 1984 | 24 | - | 278 | 1144 | - | 1446 | 1547 | 211 | 1758 |  |
| 1985 | 39 | - | 197 | 1122 | - | 1358 | 1441 | 236 | 1677 |  |
| 1986 | 26 | - | 276 | 1389 | - | 1691 | 1810 | 268 | 2078 |  |
| 1987 | 68 | - | 435 | 1419 | - | 1922 | 1958 | 314 | 2272 |  |
| 1988 | 90 | - | 584 | 1654 | - | 2328 | 2458 | 377 | 2835 |  |
| 1989 | 89 | - | 448 | 1712 | - | 2249 | 2358 | 384 | 2742 |  |
| 1990 | 82 | - | N/A | 1891 | 2 | 1977 | 2593 | 392 | 2985 |  |
| 1991 | 57 | - | 251 | 1326 | - | 1634 | 1848 | 335 | 2183 |  |
| 1992 | 25 | - | 419 | 1110 | 14 | 1568 | 1624 | 258 | 1882 |  |
| 1993 | 56 | - | 284 | 1080 | 24 | 1444 | 1417 | 197 | 1614 |  |
| 1994 | 10 | - | 277 | 998 | - | 1285 | 1156 | 248 | 1404 |  |
| 1995 | 13 | - | 288 | 857 | - | 1158 | 1031 | 216 | 1247 |  |
| 1996 | 4 | - | 279 | 855 | - | 1138 | 1044 | 222 | 1266 |  |
| 1997 | 6 | - | 329 | 1038 | 1 | 1374 | 1323 | 260 | 1583 |  |
| 1998 | 22 | - | 327 | 892 | 1 | 1242 | 1131 | 215 | 1346 |  |
| 1999 | 12 | - | 194 | 947 | - | 1153 | 1299 | 244 | 1543 |  |
| 2000 | 4 | - | 360 | 926 | <1 | 1290 | 1281 | 345 | 1625 |  |


| $\begin{aligned} & \frac{1}{\pi} \\ & \underset{\sim}{\sim} \end{aligned}$ | $\begin{aligned} & \frac{\varepsilon}{工} \\ & \frac{0}{60} \\ & \stackrel{\sim}{\infty} \end{aligned}$ |  |  |  | $\begin{aligned} & \frac{\curvearrowleft}{む} \\ & \frac{1}{\dagger} \end{aligned}$ |  |  |  |  | $\begin{aligned} & \stackrel{*}{*} \\ & \stackrel{y}{n} \\ & \stackrel{n}{0} \\ & \stackrel{0}{0} \\ & \stackrel{H}{0} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 12 | － | 303 | 797 | － | 1112 | 1106 | 204 | 1310 |  |
| 2002 | 27 | － | 242 | 978 | ＜1 | 1247 | 1257 | 215 | 1472 |  |
| 2003 | 39 | － | 216 | 985 | － | 1240 | 1277 | 110 | 1387 |  |
| 2004 | 46 | － | 184 | 912 | － | 1142 | 1212 | 126 | 1337 |  |
| 2005 | 48 | － | 198 | 887 | － | 1133 | 1203 | 117 | 1319 |  |
| 2006 | 52 | － | 223 | 965 | ＜1 | 1239 | 1313 | 97 | 1411 |  |
| 2007 | 84 | － | 202 | 680 | － | 966 | 1003 | 143 | 1146 |  |
| 2008 | 66 | － | 148 | 679 | － | 893 | 976 | 135 | 1112 |  |
| 2009 | 53 | 2 | 191 | 731 | － | 977 | 923 | 101 | 1024 |  |
| 2010 | 51 | 2 | 227 | 843 | － | 1123 | 1092 | 116 | 1208 |  |
| 2011 | 141 | 3 | 274 | 936 | － | 1354 | 1334 | 83 | 1417 |  |
| 2012 | 134 | 2 | 224 | 1004 | ＜1 | 1364 | 1366 | 126 | 1492 | 448 |
| 2013 | 97 | 1 | 221 | 1041 | － | 1360 | 1351 | 121 | 1472 | 351 |
| 2014 | 41 | － | 323 | 976 | － | 1340 | 1341 | 149 | 1490 | 1133 |
| 2015 | 111 | 1 | 224 | 912 | 1 | 1249 | 1246 | 178 | 1424 | 1276 |
| 2016 | 145 | ＜ 1 | 204 | 1430 | － | 1780 | 1777 | 235 | 2013 | 618 |
| 2017 | 151 | ＜1 | 153 | 1605 | 1 | 1911 | 1915 | 213 | 2128 | 821 |
| 2018 | 143 | 3 | 118 | 1377 | 3 | 1644 | 1644 | 236 | 1880 | 633 |
| 2019 | 73 | 2 | 97 | 1351 | ＜1 | 1523 | 1520 | 204 | 1725 | 366 |
| 2020＾ | 73 | 1 | 79 | 1122 | － | 1276 | 1275 | 98 | 1373 | 514 |
| 2021＾ | 107 | 1 | 90 | 1129 | － | 1327 | 1331 | 72 | 1403 | 211 |

＊Estimated by the working group．
${ }^{* *}$ Migration correction（ $15 \%$ of the mature population caught in Quarter 1 in Division 7．d）added to stock．
＊＊＊Discard estimated by the working group，including discards from the migration correction．
${ }^{\wedge}$ Preliminary official landings．

Table 25.2. Plaice in 7.e. Tuning fleet data available. Not all years and ages as shown here are used.

| ple.27.7e WGCSE 2022 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 103 |  |  |  |  |  |  |
| FSP-7e |  |  |  |  |  |  |
| 20032021 |  |  |  |  |  |  |
| 110.750 .80 |  |  |  |  |  |  |
| 127 |  |  |  |  |  |  |
| 1 | 0.0209264878 | 0.3436560706 | 0.343947876 | 0.2157155878 | 0.0410732591 |  |
|  | 0.0419902913 | 0.0509084187 | 0.0337008819 | 0.0219208295 | 0.002110558 |  |
|  | 0.0009455344 | 0.0004947213 | $7.72432 \mathrm{e}-05$ | $3.546 \mathrm{e}-05$ | 0 |  |
|  | $3.546 \mathrm{e}-05$ | 0 | $3.546 \mathrm{e}-05$ | 0.000191009 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.0066803368 | 0.2116598105 | 0.8405063532 | 0.1704899676 | 0.2827197449 |  |
|  | 0.0296774344 | 0.0182410083 | 0.0461270347 | 0.0114017102 | 0.0026218442 |  |
|  | 0.0008434715 | 0.0001982425 | 0.0001604926 | 0 | 0.0001907966 | 0 |
|  | 0 | 0 | 0.0001228108 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.0084930716 | 0.3271099173 | 0.4255951803 | 0.2404409927 | 0.0900371664 |  |
|  | 0.0395287705 | 0.0127361504 | 0.0174592138 | 0.0371790541 | 0.0070178609 |  |
|  | 0.0043464537 | 0 | 0 | 0 | 0.0006865187 |  |
|  | 0.0006011272 | 0.0005708894 | 0 | 0.0005994339 | 0.0014134667 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.0264706605 | 0.6226160902 | 0.4216897498 | 0.1859126341 | 0.099837907 |  |
|  | 0.0442377935 | 0.0213837161 | 0.0045703626 | 0.0063647949 | 0.0140975761 |  |
|  | 0.0015007319 | 0.0043230363 | 0 | 0 | 0.0005854935 | 0 |
|  | 0 | 0.0006888159 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.117014537 | 0.2742350811 | 0.1567513605 | 0.0653599832 |  |
|  | 0.026616889 | 0.008325896 | 0.0058923584 | 0.0054585815 | 0.005636939 |  |
|  | 0.0019316523 | 0.002303838 | 0.0022915293 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.0017564806 | 0.4978930589 | 0.2151254659 | 0.1256971005 | 0.0317951547 |  |
|  | 0.0192553747 | 0.0153145058 | 0.0046815159 | 0.001905926 | 0.0005279613 |  |
|  | 0.000223455 | 0.0001864931 | $7.71115 \mathrm{e}-05$ | $2.43906 \mathrm{e}-05$ | $2.83303 \mathrm{e}-05$ |  |
|  | $2.43906 \mathrm{e}-05$ | 0 | $2.83303 \mathrm{e}-05$ | 2.83303e-05 | $2.83303 \mathrm{e}-05$ | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.0211943046 | 0.4353288543 | 0.4422219627 | 0.1528394796 | 0.0598128665 |  |
|  | 0.0331030404 | 0.0226941756 | 0.0079648565 | 0.0033431925 | 0.0013018772 | 0 |
|  | 0.0026037544 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0.0647388571 | 0.7402133171 | 0.5829812879 | 0.3845761643 | 0.0479189382 |  |
|  | 0.0415029509 | 0.0119952249 | 0.0061701869 | 0.0023009922 | 0.0047470993 |  |
|  | 0.0011504961 | 0.006622702 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.0037734254 | 1.0381082519 | 0.8021298636 | 0.3142975263 | 0.1105006712 |  |
|  | 0.0103245315 | 0.018145561 | 0.0132876218 | 0.0020727868 | 0.0020056082 |  |
|  | 0.0016435966 | 9.86201e-05 | $6.33593 e-05$ | 0.0029840585 | 0 |  |
|  | 0.0011857569 | 0 | 3.52608e-05 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.0496877714 | 0.3213266702 | 1.2434797508 | 0.5819533643 | 0.1364606529 |  |
|  | 0.1347918963 | 0.0121371085 | 0.0144254043 | 0.0115270917 | 0.0025913556 |  |
|  | 0.0051989993 | 0.0049471333 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.2266890224 | 1.4641865697 | 1.2671043524 | 0.4403084123 |  |
|  | 0.2032552981 | 0.0755061094 | 0.0275649015 | 0.0077368789 | 0.0037886636 | 0 |
|  | 0.0012684957 | 0.0012684957 | 0 | 0 | 0 | 0 |
|  | 0.0012684957 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0.03548275 | 1.319797284 | 1.6647790616 | 2.1308293497 | 0.8328568122 |  |
|  | 0.6237114226 | 0.1572964824 | 0.0350538578 | 0.0371484109 | 0 |  |
|  | 0.0035336667 | 0.00265025 | 0 | 0 | 0.0011778889 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.0016784148 | 0.8297622663 | 1.257130266 | 0.9233582109 | 1.0196856517 |  |
|  | 0.5143815562 | 0.1222225484 | 0.0129019851 | 0.0566548058 | 0.0023922088 |  |
|  | 0.0093314682 | 4.40921e-05 | $2.2046 \mathrm{e}-05$ | $2.2046 \mathrm{e}-05$ | 0 |  |
|  | $2.2046 \mathrm{e}-05$ | 0 | $2.2046 \mathrm{e}-05$ | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.0033568296 | 0.3887210579 | 1.50073462 | 0.8157445965 | 0.3896522219 |  |
|  | 0.3426977969 | 0.2351356472 | 0.0186050167 | 0.056730528 | 0 |  |
|  | 0.0098937394 | 0 | 0.0028762403 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.0083920741 | 0.2734688532 | 0.5985856837 | 1.5809798468 | 0.3446727423 |  |
|  | 0.4536494278 | 0.2886734237 | 0.0563499887 | 0.1485529321 | 0.013814645 |  |
|  | 0.0264081583 | 0.0106433051 | 0.0054248655 | 0 | 0.0033568296 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.2730774994 | 0.765289223 | 0.9806866884 | 0.6563497784 |  |
|  | 0.1850327432 | 0.1112903326 | 0.0873483927 | 0.0302633863 | 0.0263229229 |  |
|  | 0.0030350849 | 0.0017983016 | 0 | 0 | 0 | 0 |


| 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 |  |
|  | 0.000373087 | 0.1741186522 | 0.4982278185 | 0.5246850208 | 0.4820598027 |  |
|  | 0.3039476624 | 0.1153367817 | 0.0630658514 | 0.0315684791 | 0.01902716 |  |
|  | 0.0124644406 | 0 | 0.0012684957 | 0.0015856197 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.7236873888 | 0.6876859008 | 0.286551445 | 0.1324310278 |  |
|  | 0.0833516837 | 0.0646730791 | 0.0146861296 | 0.0183201245 | 0.0088404566 |  |
|  | 0.0073055273 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.00211269 | 0.9929342784 | 1.7363595162 | 1.1438702468 | 0.1841311612 |  |
|  | 0.0630358958 | 0.0510755065 | 0.0349013699 | 0.0058015804 | 0.008438062 |  |
|  | 0.0073611358 | 0.0031734158 | 0.0031734158 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
|  | 0 | 0 | 0 | 0 | 0 |  |
| Q1SWBeam |  |  |  |  |  |  |
| 20062021 |  |  |  |  |  |  |
| $\begin{array}{llll}1 & 1 & 0 & 0.25 \\ 1 & 27 & \end{array}$ |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
| 1 | 1.84355 | 39.0324 | 28.978 | 22.789 | 6.4116 |  |
|  | 2.0366 | 0.2017 | 0.1706 | 0.3412 | 0.64363 | 0 |
|  | 0.42633 | 0.22583 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0.86782 | 16.0343 | 35.474 | 17.601 | 4.9816 |  |
|  | 4.1461 | 1.6719 | 3.5545 | 0.2503 | 4.42522 |  |
|  | 0.2503 | 2.73176 | 1.43235 | 0.2503 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.78546 | 34.0493 | 14.432 | 5.388 | 7.7622 |  |
|  | 1.1251 | 1.4744 | 2.178 | 1.979 | 0 |  |
|  | 0.87797 | 0.12102 | 0.18772 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 1.54002 | 60.3357 | 50.609 | 16.203 | 15.6705 |  |
|  | 4.8047 | 4.7493 | 0.4567 | 0.2861 | 0.45666 |  |
|  | 0.45666 | 0 | 1.86925 | 0 | 2.47554 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 1.15633 | 46.939 | 38.568 | 27.266 | 3.2952 |  |
|  | 4.1844 | 6.4977 | 0.8659 | 0.3754 | 0 | 0 |
|  | 0.18772 | 0 | 4.58514 | 0 | 0 | 0 |
|  | 0 | 0.14928 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 1.80958 | 59.7233 | 106.793 | 41.826 | 7.3508 |  |
|  | 6.3969 | 4.5944 | 0.4679 | 1.5832 | 0.11377 |  |
|  | 0.35757 | 0 | 0.11919 | 0.11919 | 0.11919 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 10.4168 | 55.348 | 34.255 | 11.0874 |  |
|  | 7.7031 | 11.5037 | 0.3472 | 3.2114 | 0 | 0 |
|  | 0.21454 | 0 | 0 | 0.1644 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0.30036 | 18.0693 | 94.898 | 71.939 | 16.1513 |  |
|  | 4.4771 | 2.337 | 1.0774 | 1.9311 | 0 | 0 |
|  | 0 | 0.13968 | 0 | 0 | 0 |  |
|  | 0.12637 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 1.01423 | 68.7637 | 155.902 |  | 70.5165 |  |
|  | 10.792 | 1.4612 | 2.9894 | 0.9387 | 0.48829 |  |
|  | 0.28101 | 0.15884 | 0.1706 | 0 | 0 | 0 |
|  | 0.15884 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 45.2386 | 48.128 | 25.168 | 37.4127 |  |
|  | 21.7209 | 5.1873 | 2.273 | 1.0775 | 2.08315 | 0 |
|  | 1.23777 | 0 | 0.18772 | 0 | 0.1976 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0.22085 | 21.6309 | 243.345 | 66.815 | $39.6987$ |  |
|  | 40.1547 | 29.5983 | 7.9856 | 17.8855 | $0$ |  |
|  | 8.59375 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 17.263 | 84.717 | 194.995 | 39.4563 |  |
|  | 37.3536 | 18.272 | 3.4646 | 6.9322 | 0.50248 |  |
|  | 0.65697 | 0 | 0.38299 | 0 | 0.2503 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 9.0674 | 40.928 | 32.98 | 49.1506 |  |
|  | 19.0937 | 13.197 | 12.3342 | 8.2191 | 1.41506 |  |
|  | 1.03906 | 0.13269 | 0 | 0.13269 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |


| 1 |  | 0 | 18.5742 | 40.702 | 52.361 | 43.0245 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 51.7778 | 21.8301 | 5.9516 | 4.1325 | 0.94572 |  |
|  |  | 0.55842 | 0 | 0 | 0.23838 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |
| 1 |  | 0 | 47.2232 | 66.046 | 34.599 | 19.2905 |  |
|  |  | 8.3288 | 6.0213 | 4.6533 | 3.5152 | 1.35155 |  |
|  |  | 3.19601 | 0.14679 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |
| 1 |  | 0 | 29.891 | 109.811 | 79.129 | 17.2037 |  |
|  |  | 6.5559 | 5.8824 | 1.7811 | 1.045 | 0.37545 |  |
|  |  | 0.18474 | 0 | 0.11942 | 0.1976 | $0$ | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |
| FSP-7e-biomass |  |  |  |  |  |  |  |
| $20032021$ |  |  |  |  |  |  |  |
| $\begin{array}{llll} 1 & 1 & 0.75 & 0.80 \end{array}$ |  |  |  |  |  |  |  |
| 127 |  |  |  |  |  |  |  |
| 1 |  | 0.005293577 | 0.121542899 | 0.153069038 | 0.109777405 | 0.02724899 |  |
|  |  | 0.030440689 | 0.039405188 | 0.023256499 | 0.021299359 | 0.003012324 |  |
|  |  | 0.001693233 | 0.000998533 | 0.000116896 | 5.67789e-05 | 0 |  |
|  |  | 5.67789e-05 | 0 | 5.67789e-05 | 0.000412429 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 |  | 0.001294171 | 0.064029538 | 0.327665102 | 0.088965348 | 0.145804905 |  |
|  |  | 0.019175154 | 0.01406247 | 0.045491867 | 0.009358115 | 0.004607872 |  |
|  |  | 0.001141269 | 0.000207496 | 0.00017526 | 0 | 0.000296993 | 0 |
|  |  | 0 | 0 | 0.000240839 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |
| 1 |  | 0.001834264 | 0.111584597 | 0.162368193 | 0.11579134 | 0.054922714 |  |
|  |  | 0.027075454 | 0.010259167 | 0.017859677 | 0.034554905 | 0.007686843 |  |
|  |  | 0.006065153 | 0 | 0 | 0 | 0.001218691 |  |
|  |  | 0.000962531 | 0.001353022 | 0 | 0.001294305 | 0.002771882 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 |  | 0.004986995 | 0.18835206 | 0.153217512 | 0.097268733 | 0.060978776 |  |
|  |  | 0.028978287 | 0.015872993 | 0.004062249 | 0.006358017 | 0.015281921 |  |
|  |  | 0.002559487 | 0.003623002 | 0 | 0 | 0.000347422 | 0 |
|  |  | 0 | 0.001102939 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |
| 1 |  |  |  |  | 0.070882913 | $0.039117126$ |  |
|  |  | 0.02017179 | 0.007948558 | 0.003676961 | $0.005215244$ | $0.005702262$ |  |
|  |  | 0.002246861 | 0.004195415 | 0.001280989 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |
| 1 |  | 0.000762555 | 0.165757719 | 0.09611746 | 0.070482203 | 0.023197206 |  |
|  |  | 0.016585008 | 0.014802254 | 0.005977404 | 0.00156398 | 0.00075609 |  |
|  |  | 0.000306786 | 0.00031644 | 0.000118364 | $3.14036 \mathrm{e}-05$ | 5.55573e-05 |  |
|  |  | $3.14036 \mathrm{e}-05$ | 0 | 5.55573e-05 | 5.55573e-05 | 5.55573e-05 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 |  | 0.004328708 | 0.154089593 | 0.197184164 | 0.080551521 | 0.040774069 |  |
|  |  | 0.030526187 | 0.018386791 | 0.007170953 | 0.005353153 | 0.00255305 | 0 |
|  |  | 0.003877076 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 |  |  |
| 1 |  | 0.007795399 | 0.222343894 | 0.24054883 | 0.200768822 | 0.032681993 |  |
|  |  | 0.028385365 | 0.011556535 | 0.006137064 | 0.002962591 | 0.00716298 |  |
|  |  | 0.002042332 | 0.003041061 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |
| 1 |  | 0.000392661 | 0.283172819 | 0.275995973 | 0.143226456 | 0.067072751 |  |
|  |  | 0.006345669 | 0.019907033 | 0.010975028 | 0.002894182 | 0.002763981 |  |
|  |  | 0.002248034 | 0.000153348 | $9.68876 \mathrm{e}-05$ | 0.004214009 | 0 |  |
|  |  | 0.001537756 | 0 | $5.646 \mathrm{e}-05$ | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 |  | 0.008807035 | 0.081378624 | 0.400587835 | 0.235678139 | 0.076449968 |  |
|  |  | 0.080259232 | 0.013262699 | 0.009394429 | 0.011571324 | 0.002465811 |  |
|  |  | 0.005859429 | 0.004732132 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |
| 1 |  | 0 | 0.049342684 | 0.401684632 | 0.466537624 | 0.206912371 |  |
|  |  | 0.108655459 | 0.054254917 | 0.020215075 | 0.010110847 | 0.003552844 |  |
|  |  | 0.001455044 | 0.001455044 | 0 | 0 | 0 | 0 |
|  |  | 0.003290293 | 0 | 0 | 0 | 0 | , |
|  |  | 0 | 0 | 0 | 0 |  |  |
| 1 |  | 0.0041401 | 0.296033996 | 0.456780968 | 0.717682874 | 0.340722206 |  |
|  |  | 0.261575403 | 0.079435747 | 0.025108143 | 0.029091589 | 0 |  |
|  |  | 0.004970024 | 0.001837137 | 0 | 0 | 0.000926584 | O |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |
| 1 |  | 0.000166693 | 0.185697599 | 0.396768066 | 0.354955458 | 0.432597481 |  |
|  |  | 0.254028547 | 0.077329369 | 0.012734295 | 0.037467198 | 0.001575026 |  |
|  |  | 0.013224445 | $7.06007 \mathrm{e}-05$ | $3.53004 \mathrm{e}-05$ | $3.53004 \mathrm{e}-05$ | 0 |  |
|  |  | $3.53004 \mathrm{e}-05$ | 0 | $3.53004 \mathrm{e}-05$ | 0 | 0 | 0 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 |  | 0.000257249 | 0.088731125 | 0.432170069 | 0.33029376 | 0.187015182 |  |
|  |  | 0.180552834 | 0.125276737 | 0.013959313 | 0.045946967 | 0 |  |
|  |  | 0.00886356 | 0 | 0.001943963 | 0 | 0 | 0 |


|  | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.000716617 | 0.063625731 | 0.168860623 | 0.550867194 | 0.152306457 |  |
|  | 0.207245338 | 0.155649866 | 0.03381067 | 0.09188589 | 0.012446502 |  |
|  | 0.019714288 | 0.010609453 | 0.005307667 | 0 | 0.003212512 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.066152 | 0.212699658 | 0.326325231 | 0.283598616 |  |
|  | 0.105983836 | 0.068842772 | 0.069157464 | 0.022306102 | 0.023142302 |  |
|  | 0.003509849 | 0.003526564 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | $3.70533 \mathrm{e}-05$ | 0.039324415 | 0.147795101 | 0.176720127 | 0.205196314 |  |
|  | 0.15664294 | 0.069133305 | 0.045915506 | 0.021835755 | 0.017404333 |  |
|  | 0.010495682 | 0 | 0.001633223 | 0.000688378 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 0.142339922 | 0.17957078 | 0.099796 | 0.063359982 |  |
|  | 0.049534859 | 0.047201656 | 0.017829374 | 0.019932615 | 0.008423159 |  |
|  | 0.009645627 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.000113987 | 0.16918976 | 0.391374814 | 0.296746707 | 0.075908191 |  |
|  | 0.032207573 | 0.033996438 | 0.02900624 | 0.006577623 | 0.008533158 |  |
|  | 0.005266963 | 0.005390231 | 0.003227407 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |

Table 25.3. Plaice in 7e. von Bertalanffy growth parameters.

| Year | $k\left(\right.$ year $\left.^{-1}\right)$ | $L_{\infty}(\mathbf{c m})$ | $t_{0}$ (years) |
| :---: | :---: | :---: | :---: |
| 2006 | 0.14 | 57.97 | -2.09 |
| 2007 | 0.20 | 52.96 | -1.28 |
| 2008 | 0.23 | 47.53 | -1.43 |
| 2009 | 0.19 | 52.38 | -1.43 |
| 2010 | 0.21 | 53.59 | -1.07 |
| 2011 | 0.15 | 58.64 | -1.43 |
| 2012 | 0.17 | 56.71 | -1.04 |
| 2013 | 0.14 | 60.02 | -1.61 |
| 2014 | 0.15 | 56.18 | -1.15 |
| 2015 | 0.12 | 62.35 | -1.66 |
| 2016 | 0.11 | 59.99 | -2.28 |
| 2017 | 0.11 | 58.88 | -1.91 |
| 2018 | 0.15 | 54.06 | -1.28 |
| 2019 | 0.12 | 56.54 | -1.98 |
| 2020 | 0.11 | 62.42 | -2.53 |
| 2021 | 0.08 | 61.58 | -3.39 |
| 2017-2021 | 0.11 | 58.51 | -2.30 |

Table 25.4. Plaice in 7e. Parameters of the commercial catch length distribution. The table shows length at first capture $L_{c}$ and mean length in the catch $L_{\text {mean }}$ above $L_{c}$.

| Year | Annual $\boldsymbol{L}_{\boldsymbol{c}}(\mathbf{c m})$ | $\boldsymbol{L}_{\boldsymbol{c}}(\mathbf{c m})$ (used for $\left.\boldsymbol{L}_{\text {mean }}\right)$ | $\boldsymbol{L}_{\text {mean }} \mathbf{( c m )}$ |
| :---: | :---: | :---: | :---: |
| 2015 | 26 | 26.4 | 32.7 |
| 2016 | 26 | 26.4 | 33.7 |
| 2017 | 27 | 26.4 | 33.0 |
| 2018 | 26 | 26.4 | 33.1 |
| 2019 | 26 | 26.4 | 34.4 |
| 2020 | 26 | 26.4 | 32.6 |
| 2021 | 27 | 26.4 | 32.3 |
| $2017-2021$ | 26.4 |  |  |

Table 25.5. Plaice in 7e. The basis for the catch options for 2021. Note that one catch option is provided for stocks in ICES data categories 3-6. The values presented here are the values presented during the working group.

| Division 7.e plaice stock |  |
| :---: | :---: |
| Previous catch advice $\mathrm{A}_{\mathrm{y}}$ (advised catch for 2022) | 1742 tonnes |
| Stock biomass trend |  |
| Index A (2020, 2021) | $0.81 \mathrm{~kg} \mathrm{hr}^{-1} \mathrm{~m}_{\text {beam }}{ }^{-1}$ |
| Index B (2017, 2018, 2019) | $1.10 \mathrm{~kg} \mathrm{hr}^{-1} \mathrm{~m}_{\text {beam }}{ }^{-1}$ |
| $r$ : Stock biomass trend (index ratio $A / B$ ) | 0.74 |
| Fishing pressure proxy |  |
| Mean catch length ( $L_{\text {mean }}=L_{\text {2021 }}$ ) | 32.3 cm |
| MSY proxy length ( $L_{F=M}$ ) | 34.4 cm |
| f: Fishing pressure proxy relative to MSY proxy ( $L_{2021} / L_{\text {F }}$ M ) | 0.94 |
| Biomass safeguard |  |
| Last index value ( $\mathrm{I}_{2021}$ ) | $1.03 \mathrm{~kg} \mathrm{hr}^{-1} \mathrm{~m}_{\text {beam }}{ }^{-1}$ |
| Index trigger value ( $l_{\text {trigger }}=l_{\text {loss }} \times 1.4$ ) | $0.39 \mathrm{~kg} \mathrm{hr}^{-1} \mathrm{~m}$ beam ${ }^{-1}$ |
| b: Index relative to trigger value, $\min \left\{\mathrm{I}_{2021} / I_{\text {trigger }}, 1\right\}$ | 1 |

Precautionary multiplier to maintain biomass above $\mathrm{B}_{\text {lim }}$ with $95 \%$ probability

| $m:$ multiplier (generic multiplier based on life history) | 0.95 |
| :--- | :--- |
| Uncertainty cap $\left(+20 \% /-30 \%\right.$ compared to $A_{y}$, only applied if $\left.b \geq 1\right)$ | Applied |
| Discard rate | $27 \%$ |
| Catch advice for 2023 and $2024\left(A_{y} \times\right.$ uncertainty cap) | 1219 tonnes |
| Projected landings corresponding to advice** | 894 tonnes |
| $\%$ advice change^ | $-30 \%$ |

Plaice in Division 7.e

| Proportion of Division 7.e stock landings caught in Division 7.e (2003-2021) | 0.91 |
| :--- | :--- |
| Catch of plaice in Division 7.e corresponding to the advice for the stock | 1104 tonnes |
| Projected landings of plaice in Division 7.e corresponding to the advice for the stock** | 809 tonnes |

* The figures in the table are rounded. Calculations were done with unrounded inputs, and computed values may not match exactly when calculated using the rounded figures in the table.
** [Advised catch for 2023] $\times$ [1 - discard rate].
${ }^{\wedge}$ Advice value for 2023 relative to the advice value for 2022 ( 1742 tonnes).


### 25.17 Figures



Figure 25.1. Plaice in 7.e. International landings and discards by country as extracted from InterCatch.


Figure 25.2. Plaice in 7.e. International landings and discards reported to InterCatch per country and fleet.


Figure 25.3. Plaice in 7.e. Landings and discards of the plaice 7.e stock disaggregated by the 7.e and the migration component from 7.d. Discard data are only available starting from 2012 for the Division 7.e.


Figure 25.4. Plaice in 7.e. Discard rates. "Fleet mean" is the mean of the ratios for all fleets which reported discards, "reported" is the proportion of reported discards in the reported catches, "weighted fleet mean" is the mean of the ratios for all fleets which reported discards weighted by the catch of the individual fleets, "raised" is the proportion of the discards as raised within InterCatch in the total catch for 7.e and "raised incl. migration" includes the catch (discards and landings) from Division 7.d used in the migration correction.


Figure 25.5. Plaice in 7.e. Landings, Discards and discard rate.


Figure 25.6. Plaice in 7.e. Age samples from InterCatch. The numbers are raised to fleet level.


Figure 25.7. Plaice in 7.e. Length samples from InterCatch. The numbers are raised to fleet level.


Figure 25.8. Plaice in 7.e. Landings age distribution.


Figure 25.9. Plaice in 7.e. Derivation of the stock and catch (landings) weights by applying a polynomial model to the raw InterCatch weights-at-age.


Figure 25.10. Plaice in 7.e. Landings and stock weights-at-age used in the assessment.


Figure 25.11. Plaice in 7.e. Scientific tuning information from the two surveys. For the UK-FSP survey, numbers and biomass are shown.


Figure 25.12. Plaice in 7.e. Scientific tuning information from the two surveys, standardised to the mean of the timeseries and cohort-wise. For the UK-FSP survey, numbers and biomass are shown.


Figure 25.13. Plaice in 7.e. Internal consistency of the two survey time-series including correlation analysis.


Figure 25.14. Plaice in Division 7.e. Plaice catch rates during FSP "Western Channel Sole and Plaice" surveys, 2003-2020 (number $\mathrm{h}^{-1} \mathrm{~m}^{\text {beam }}{ }^{-1}$ ). Open circles: FV Nellie and FV Carhelmar tows; filled black circles: FV Lady T Emiel tows. Please note that 2021 numbers are not to scale. Source: Burt et al. (2021, 2022).


Figure 25.15. UK commercial LPUE time-series. LPUE values are only shown for historical reasons but were not used in the assessment.


Figure 25.16. Plaice in 7.e. Age-length keys. The red circles represent data from sampling of the UK commercial catches and blue/green circles represent the fish aged from the Q1SWBeam survey. Solid black curves indicate annual fits of a von Bertalanffy growth function, dashed lines a fit to the combined data from the last five years (2017-2021).


Figure 25.17. Plaice in 7.e. Comparison of historical catch and ICES advice.


Figure 25.18. Plaice in 7.e. The biomass index used in the rfb rule. The biomass index is based on the UK-FSP surey and includes ages 2-8.


Figure 25.19. Plaice in 7.e. Total international length frequencies for 2014-2021 as raised within InterCatch for landings and discards including Length of first capture (Lc, calculated as first length class where the abundance is bigger or equal to half of maximum abundance) and mean length in the catch (Lmean, mean length above Lc).

Length indicator


Figure 25.20. Plaice in 7.e. Mean catch length in comparison to the MSY proxy reference length $\mathrm{L}_{\mathrm{F}=\mathrm{M}}$.


Figure 25.21. Plaice in 7.e. Summary of the legacy XSA landings-only assessment. Reference points in this assessment are from ICES WKMSYREF4 (ICES, 2016). Please note that this assessment is not used for advice purposes.


$\log$ residuals

- 0.01
- 0.10

○ 0.50
○ 1.00
$\bigcirc 2.00$

- negative
- positive




Figure 25.22. Plaice in 7.e. Residuals of the legacy XSA landings-only assessment. Please note that this assessment is not used for advice purposes.


Figure 25.23. Plaice in 7.e. Retrospective analysis of the legacy XSA landings-only assessment. Please note that this assessment is not used for advice purposes.


Figure 25.24. Plaice in 7.e. Summary of an exploratory SAM assessment. Please note that this assessment is not used for advice purposes.

## a) retro




c) residuals

e) jitter runs



b) leave-out



d) process residuals

f) simstudy




Figure 25.25. Plaice in 7.e. Model diagnostics of the exploratory SAM assessment. Please note that this assessment is not used for advice purposes.

## 26 Plaice in Divisions 7.f-g (Celtic Sea)

### 26.1 Type of assessment in 2022

In February 2022 the stock was benchmarked, and the State-Space Assessment Model (SAM) was suggested for the stock to be assessed as the Category 1. However, WGCSE rejected this approach due uncertainties with natural mortality and doubts in model configuration. A SPiCT model although it converged, there was a very large uncertainty and unsatisfactory diagnostics. WGCSE concluded that the stock to be assessed as the category 3 with the application of ICES technical guidance for harvest control rules the rfb was applied. The advice is based on the recent advised catches, multiplied by the ratio of the mean of the last two index values (index A) and the mean of the three preceding values (index B), a ratio of observed mean length in the catch relative to the target mean length, a biomass safeguard, and a precautionary multiplier. Mean catch length (Lmean) in recent years was below of the MSY proxy length (Lf=m) reference point, and biomass index was below the index trigger value. Because of this the $-30 \%$ uncertainty cap was not applied.

## ICES advice applicable to 2021

Based on the ICES approach for data-limited stocks, ICES advises that catches in 2021 should be no more than 1911 tonnes.

## General stock description and management units

A TAC is allocated to ICES areas 7.f-g which corresponds to the stock area.

## Management applicable to 2021

TACs and quotas set for 2021 (source COUNCIL REGULATION (EU) No 1239/2021)
Species: Plaice Pleuronectes platessa, Zone: 7.f and 7.g (PLE/7FG.)
Belgium ..... 360
France ..... 648
Ireland ..... 240
United Kingdom ..... 480
Total EU ..... 1249
Total TAC ..... 1911

TACs and quotas set for 2022 (source COUNCIL REGULATION (EU) No 109/2022)
Species: Plaice Pleuronectes platessa, Zone: 7.f and 7.g (PLE/7FG.)

| Belgium | 379 |
| :--- | :--- |
| France | 686 |
| Ireland | 105 |
| Total EU | 1170 |
| United Kingdom | 441 |
| Total TAC | 1735 |

## Fishery in 2021

As usual the main fishery was concentrated on the Trevose Head ground off the north Cornish coast and around Land's End with most of the catch being taken in the eastern part of the area (Figure 26.1). Plaice was harvested throughout the year, with the highest amount of fish landed in Q3 and the lowest in Q1. The fleets harvesting plaice in the Celtic Sea primarily involved vessels from Belgium, France, Ireland and the UK with negligible amounts taken by Netherlands and Spain. In 2021 Belgium reported $56.2 \% \%$ of the landings, France $10.9 \%$, Ireland $22.4 \%$ and the UK $10.3 \%$. The contribution of individual countries to total landings was similar to 20132020. The Working Group estimated that total international landings for 2021 were 468 t with the reconstructed total catch (including discards) being estimated as 846 t . It is $56.7 \%$ lower than the TAC of 1911 t , which also included discards (Table 26.1). Discards represented some $44.6 \%$ of the catch.

Most of the catch ( $80.1 \%$ ) were taken by beam trawlers, and $17.9 \%$; by bottom otter trawlers. Other gears accounted for $2.0 \%$. Effort and lpue of fishing fleets are presented in Tables 26.226.4 .

### 26.2 Data

## Landings

International catch data are collated on the ICES InterCatch platform (https://intercatch.ices.dk). All landings are reported and recreational catch is supposed to be negligible.

## Discards

Discarding with this fishery is considered high (Table 26.1). During the 2022 working group, discard information was made available as annual summaries for Belgium and Ireland, and on quarterly basis for France and UK. If sampling information was not available, discard estimates were calculated based on the discard rates of similar fleets in terms of gear and time (quarter /annual). WG estimates of discards, show a steady increase from 2004 (when estimates started) to a peak in 2013, after which discard levels have remained variable but high, with discard estimates regularly exceeding landings, where they were lower than landings. Data from national discard sampling programmes are summarised in Figures 26.2-26.4.

## Biological information

Age compositions for 2021 were supplied at an annual or quarterly disaggregation for Belgium, Ireland, and UK(E+W), providing a sampling coverage of $81.6 \%$ for the total landings and $77.4 \%$ for the discards.

Figure 26.4 compares the landings and discard numbers-at-age (where data are available). A strong recruitment cohort that appeared first in 2012 as 2 y.o., in 2015 attained the age of 5 y.o. and began to predominate in landings, being still important in 2017 as 7 y.o fish. The next moderately strong generation (2 y.o. in 2015) in 2017 represented important part of both landings and discards being the most abundant age group in 2018 when 5 y.o. (Figure 26.4). Another reasonably abundant generation ( 3 y.o. in 2018) in 2019 was the most abundant group In recent years, the bulk of catches was represented by fish born in 2017-2018 (2-3 y.o in 2020 and 3-4 y.o. in 2021). Numbers- and weights-at-age for landings, discards and the stock used in the assessment are presented in Tables 26.5 and 26.7.

## Landings weight-at-age

Historically, landings weights-at-age were constructed by fitting a quadratic smoother through the aggregated catch weights for each year. WKFLAT (2011) decided not to continue with this approach following concerns raised by WGCSE that poor fits of the quadratic smoothing curve were resulting in the youngest ages being estimated to have heavier weights than adjacent older ages. WKFLAT (2011) rejected the use of the polynomial smoother for weights-at-age and suggested that raw landings weights are used in future. Raw data back to 1995 were obtained by WKFLAT (2011) and used to update the catch weights and stock weights files (Tables 26.6 and 26.9). Now weights-at-age of landings are coming from landings of Belgium, Ireland and United Kingdom age and integral values were estimated using ICES InterCatch web-based system.

## Discard weight-at-age

Discard length and weight-at-age raw data were available for Belgium, Ireland and United Kingdom and integral values were estimated using ICES InterCatch web-based system (Tables 26.7 and 26.8).

## Stock weight-at-age

Where discard estimates were available from 2004 onwards, a revised set of stock weights-at-age were calculated. The stock weights were derived from the total international landings weights-at-age and the discard weights-at-age averaged by numbers-at-age from the respective datasets. Prior to 2004, a revised set of stock weights-at-age based on international landings data was produced. These new values were based on collected weight data with a SOP correction (Table 26.9).

## Natural mortality and maturity

Estimates of natural mortality and maturity were not used in this assessment.

## Surveys

Indices of abundance from the UK(E\&W)-BTS-Q3 beam trawl survey in 7.f and the Irish IBTS survey (IGFS-WIBTS-Q4) in 7.g are presented in Table 26.10. Both surveys show consistent trends
of the stock increases and decreases (Figure 26.5). The UK(E\&W)-BTS-Q3 started in 1988 and was always used for tuning the AP model. The Irish Celtic Explorer IBTS survey (IGFS-WIBTS-Q4) time-series started in 2003 and was not used in earlier years. Both survey time-series were used for the stock trends-based advice in the years 2015, 2016 and 2017 and for SPiCT in 2018-2020.

In 2022, before applying rfb rule the UK(E\&W)-BTS-Q3 survey and IGFS-WIBTS-Q4 survey total biomass indices were standardised and compared. The UK BTS was chosen for the assessment as carried out by a proper gear efficiently catching flatfish occurs and across the area of maximum abundance (Figure26.5).

## Commercial landings per unit of effort

Commercial indices of abundance from the different fisheries provide contradictory trends (Figures 26.6 and 26.7), due to varying discarding practices in the fleets, and the discarding of fish above minimum conservation reference size. Therefore, these LPUEs, regardless their precision and objectiveness, could not be considered as proxies for adult fish abundance. However, in 2018 and particularly in 2019-2021 the situation began to return to normal when most of fish of commercial size was retained (Figure 26.8).

Belgium beam-trawling fleet takes the largest portion of catch. Therefore, it potentially might be the most reliable source of LPUE data to be used in an age-structured model. However, the Belgian landings and effort data are influenced by policy decisions, particularly often changing limits of how much plaice might be taken per fishing days depending on season, year and boat size. The resulting LPUE therefore should only be used for indicative purposes and be considered qualitatively (Nimmegeers et al., 2021).

## Other relevant data

There were no early closures of the fishery for plaice in 2021. Misreporting is not considered to be a problem in this stock. Recent research on discard survival in the English Channel has indicated that discard mortality of adult plaice captured by beam trawl varied with season, fish size and other factors like vessel type (Revill et al., 2013; Depestele et al., 2014; Uhlmann et al., 2016 a,b). Therefore, significant amounts ( 4 to $93 \%$, mostly $<50 \%$ in Belgian beam trawlers and mean $48 \%$ in French beam trawlers) might survive discarding which has been confirmed by several (315) days of observations in captivity (Depestele et al., 2014; Uhlmann et al., 2016 a). The survival estimate for the UK otter trawl fishery in the Western Channel was $47-63 \%$ and for the trammel net fishery $71-72 \%$. The discard survival was also estimated as $19-20 \%$ for the North Sea UK otter trawl fishery and $4-15 \%$ in the Western Channel UK beam trawl fishery (Catchpole et al., 2015). Smaller undersized plaice that represent the bulk of discards are likely to have relatively higher mortality as with other flatfish species (review: Hendrikson, Nies, 2007). Generally, discard survival is expected to be $\sim 40 \%$ (Catchpole et al., 2015). There is no formal mixed-fishery analysis for this area, but plaice in 7.fg is considered to be primarily a bycatch of the targeted sole fishery, so changes in effort in the directed sole fishery as well as multiannual management measures (EU 2019) will impact fishing mortality on plaice.

### 26.3 Stock assessment

## Assessment model

WKFLAT (2011) agreed that the model that will be used as a temporary basis for the assessment and provision of advice for the Celtic Sea plaice is AP model (Aarts and Poos, 2009). This was
selected on the basis that it was the only model available to WKFLAT which reconstructs the historic discarding rates (derived from the survey dataseries).

WKFLAT (2011) concluded that:
1 ) Due to the change in estimated fishing mortality when discards are included within the model fit, discards should be retained within the assessment model structure.

2 ) Given that the time-series of discard data, to which the models are fitted, is short and that, consequently, there are likely to be changes in the management estimates as discard data are added in subsequent years, no definitive model structure can be recommended at this stage in the development process.
3 ) The most flexible of the models TVS_PTVS should be used as the basis for advice; in terms of relative changes in estimated total fishing mortality and biomass.
4 ) The other two models which provide similar structures should continue to be fitted at the WG to provide sensitivity comparisons.
5 ) As the dataseries are extended, a final model selection can be then determined.

In 2013, no assessment was presented for this stock given that the "preferred" Aarts and Poos (2009) model failed to converge and other model variants could not provide realistic representations of observed landings and discards. Consequently, WGCSE 2013 decided to avoid the use of the "preferred" TV_PTVS AP model variant and instead focus on assessing the stock using trends derived from the fishery-independent $\mathrm{UK}(\mathrm{E} \& W)$ beam trawl survey. Trends derived from the UK(E\&W) beam trawl survey were selected for the basis of advice given that this survey most appropriately covered the spatial extent of the stock and well represented the mean age (25) landed in the fishery. The UK $(E \& W)$ beam trawl survey was used to infer trends in recruitment, stock size (spawning-stock biomass) and fishing mortality.

In 2014 corrected TV_PTVS Aarts and Poos (2009) model converged and produced realistic results and confirmed conclusions derived in 2013 from the fishery-independent UK(E\&W) beam trawl survey. In 2015-2017 all three model variants converged, but only of the "preferred" TV_PTVS AP variant provided estimations consistent with the previous run, observed catches and landings. However, trends of both UK(E\&W)-BTS-Q3 beam trawl and IGFS-WIBTS-Q4 surveys on one hand (Table 26.10) and data on LPUEs of commercial fleets (Tables 26.2-26.4) produced conflicting signal that resulted in asymmetrical distribution of residuals. Because of this, the ICES stock advice was based on both surveys' CPUE trends.

Independently of WGCSE, the stock status was explored in 2015 by WKLIFE using a biomass dynamic model (SPiCT) (ICES, 2016 a). As discard data were not available prior 2004, the group approximated the total catch values from 1977 to 2003. An adjustment was made to the data by applying the 2004 discard ratio back in time (landings prior to 2004 were multiplied by $\mathrm{K}=1.54$ ). These total catch data were combined with CPUE trends of both surveys expressed in two meanstandardized biomass index series of +3 -year-old plaice, which were considered to reflect "exploitable biomass" for this stock.

Results of modelling were found to be sensitive to truncating the catch to ensure $100 \%$ overlap between the survey and catch time-series. In this case, truncation lead to a $\sim 60 \%$ increase in $B_{\text {MSY }}$ and $\sim 30 \%$ decrease in FMSY, whereas CVs were hugely increased (by $\sim 200 \%$ and $\sim 75 \%$ respectively). Therefore, the time-series was not truncated. Estimation of the observation error corresponding to the catch $(\beta)$ and survey $(\alpha)$ was tried, but the model did not converge when trying to estimate both of these, so $\alpha$ was fixed at 1 , while $\beta$ was estimated. Under all these assumptions the results indicated current stock status (2015) to be well above the biomass reference point 0.5 $B_{\text {msy, }}$ and F (2015) to be well be FMSY (ICES, 2016a).

In 2017, the ICES framework for category 3.2 stocks was applied (ICES, 2012; 2016 b-d). As the previous ICES advice used both catch/landings and biomass index series, the stock was investigated by applying SPiCT. The SPiCT results were chosen to support the basis for advice using comparison of the two latest biomass index ( $\mathrm{B} / \mathrm{BMSY}^{\text {) }}$ ) values (index A ) with the three preceding values (index B), multiplied by the recent advised catch. The same approach was used later in 2018-2021.

In February 2022, the stock was benchmarked, and the State-Space Assessment Model (SAM) was suggested for the stock to be assessed as the Category 1. The stock assessment from the benchmark is available from stockassessment.org:

## Library(stockassessment)

fit_online <- fitfromweb("plaice7fg_001") (Figure 26.9)
The working group in May 2022 found two important issues with applied parameters.
The model configuration assumed that the F-at-age lognormal random walks are not correlated across ages which are not explained or justified in the (draft) benchmark report. In practice it led to the fishing mortality trajectories being very smooth throughout the history (Figure 26.9).

The correlation might be included with an AR1 process (the default for SAM): conf_F <- fit\$conf

```
conf_F$corFlag <- 2
```

pars_F <- stockassessment::defpar(fit\$data, conf_F)
fit_F <- sam.fit(data $=$ fit\$data, conf $=$ conf_F, parameters $=$ pars_F)
and compare to the benchmark model (Figure 26.10):
$\operatorname{par}(\operatorname{mar}=c(2,4.5,0.5,0.5))$
$\operatorname{plot}(\mathrm{c}($ benchmark $=$ fit, correlatedF $=$ fit_F $))$

It would lead to substantial changes in SSB and F in recent years (Figure 26.10). Also, it causes a strong retrospective pattern (Figure 26.11) with high Mohn's Rho.
mohn(retro_F)

| \#\# | R(age 1) | SSB | Fbar(3-6) |
| :--- | :--- | :--- | :--- |
| \#\# | -0.1963765 | 0.5278464 | -0.3554554 |

In the benchmark model, the recruitment in the assessment is modelled as a constant, which also was not justified. However, removing this assumption only slightly changes the model output.

On another hand there was a notable influence coming from the choice of natural mortality in the estimation of reference points. The WKNSCS Benchmark 2022 meeting reviewed a model incorporating natural mortality estimates taken from the plaice assessment in the Irish Sea, Division 27.7a (ple.27.7a). A second variant was subsequently established (Ple.7fg_disc on Stockassessment.org), using natural mortality estimates used in the assessment of plaice in the English Channel (ple.27.7d), calculated using the Gislason equation, which were notably higher. These were used as data were not appropriate (lacking older ages) for the Celtic Sea ( 27.7 fg ) stock. In both model variants, natural mortality was specified for each of the modelled age classes ( 1 to 10 year olds) and fixed across the full time-series.

For the SAM model seen by the benchmark process WKNSCS (low natural mortality), the MSY $B_{\text {trigger }}$ reference point is notably higher, while $B_{l i m}$ and $B_{\text {trigger, }}$ lower than the alternative model. F reference points are all lower (Table 26.11).

The premise for this is that with the lower natural mortality values, taken from the English Channel in the original run, a greater proportion of total mortality in the SAM assessment model is being attributed to fishing pressure and therefore suggest that $F$ has been relatively high and has been limiting stock development. It suggests if fishing pressure were reduced a large development of the stock would result, and a high yield even at a much lower $F$ than is currently observed.

A SPiCT model was also explored, although it converged, there was a very large uncertainty and unsatisfactory diagnostics. Therefore, WGCSE concluded that the stock to be assessed as the category 3 with the application of ICES technical guidance for harvest control rules the rfb was applied. The rfb rule is an empirical harvest control rule meant to be applied to empirical data and not estimates from a population model.
The rfb rule is defined as:

$$
A_{y+1}=A_{y} r f b m
$$

where $A_{y+1}$ is the new advised catch for year $y+1, A_{y}$ the previously advised catch, $r$ a biomass ratio following the trend in a biomass index, $f$ a fishing pressure proxy using catch length data, $b$ a biomass safeguard protecting against low stock size, and $m$ a multiplier ensuring long-term precautionary exploitation (ICES, 2022).
The rfb rule includes a conditional and asymmetric uncertainty cap restricting changes in $A_{y+1}$ relative to $A_{y}$ to $+20 \%$ and $-30 \%$, but is only implemented if $b=1$. Furthermore, the rfb rule provides biennial catch advice, i.e. once set, the advice is kept for two years (ICES, 2022).

Growth rates were estimated using data from Belgium commercial samples and both UKBTS and IGFS surveys (Figure 26.12). Mean coefficient $k$ of von Bertalanffy for 2016-2020 equation was 0.0985 and as $\mathrm{k}<0.2$ the rfb rule method 2.1 with $\mathrm{m}=0.95$ should be applied (ICES, 2022). The length structure was obtained from InterCatch for a period of 2018-2021, age data were obtained from UK and Ireland research surveys and catches of Belgian beam trawlers for a period 20002021. CPUE of UKBTS survey for a period of 1988-2021 expressed in total catch (kg) per 100 km of towed distance was used as the index of abundance. Lc was stable during recent years being $26.3-30.1 \mathrm{~cm}$, below MSY proxy length ( $\mathrm{L}_{\mathrm{F}=\mathrm{m})}=30.2 \mathrm{~cm}$ (Figure 26.13). As the recent advices were very different from actual catches, the rfb harvest control rule was based on catches as this rule is meant to adjust realised catches influencing the stock (Fischer et al., 2020).

The advice is based on the recent advised catches, multiplied by the ratio of the mean of the last two index values (index A) and the mean of the three preceding values (index B), a ratio of observed mean length in the catch relative to the target mean length, a biomass safeguard, and a precautionary multiplier. The stability clause was not applied because the recent biomass index value was below MSY Btrigger. The discard rate (average 2019-2021) was $38 \%$. The resulted advice is provided in the table below.

| Mean catch last three years ( $\mathrm{C}_{\mathbf{y}}$ ) | 793 tonnes |
| :---: | :---: |
| Biomass index trend |  |
| Index A (2020-2021) | $58.9 \mathrm{~kg} / 100 \mathrm{~km}$ |
| Index B (2017-2019) | $89.9 \mathrm{~kg} / 100 \mathrm{~km}$ |
| r: Stock biomass trend (index ratio $A / B$ ) | 0.655 |
| Fishing pressure proxy |  |
| Mean catch length ( $L_{\text {mean }}=L_{\text {2021 }}$ ) | 26.3 cm |
| MSY proxy length ( $\mathrm{L}_{\mathrm{F}=\mathrm{M}}$ ) | 30.2 cm |
| f: Fishing pressure proxy relative to MSY proxy ( $\mathrm{L}_{2021 / \mathrm{LF}=\mathrm{M}}$ ) | 0.937 |
| Biomass safeguard |  |
| Last index value ( $\mathrm{I}_{2021}$ ) | 49.3 kg / 100 km |
| Index trigger value ( $\left.l_{\text {trigger }}=l_{\text {loss }} \times 1.4\right)$ | 57.0 kg / 100 km |
| b: index relative to trigger value, $\min \left\{I_{2021} / I_{\text {trigerer }}, 1\right\}$ | 0.87 |
| Precautionary multiplier to maintain biomass above $\mathrm{B}_{\text {lim }}$ with $95 \%$ probability |  |
| m : multiplier (generic multiplier based on life history) | 0.95 |
| Uncertainty cap (+20\%/-30\% compared to $A_{y}$, only applied if $b \geq 1$ ) | Not applied |
| Discard rate | 38\% |
| Catch advice for 2023 and 2024 [ $\mathrm{C}_{\mathrm{y}} \times \mathrm{r} \times \mathrm{fx} \times \mathrm{bx} \mathrm{m}$ ] | 402 tonnes |
| \% advice change** | -73\% |

## State of the stock

Index of the stock biomass is estimated to have been increasing between 2005 and 2015 and from 2016 went into decline descending by 2021 to values close to the lowest observed (Figure 26.5).

### 26.4 Short-term projections

The short-term projections for the stock are not established.

### 26.5 Precautionary approach reference points

Mean catch length (Lmean) in recent years was below of the MSY proxy length ( $\mathrm{Lf}=\mathrm{m}$ ) reference point established as 30.2 cm TL, and last index value was below the index trigger value established as $57.0 \mathrm{~kg} / 100 \mathrm{~km}$. Because of this the $-30 \%$ uncertainty cap was not applied.

| Year | Lmean, cm | $\mathbf{L c}, \mathbf{c m}$ | $\mathbf{L}_{\mathrm{F}=\mathrm{m}, \mathrm{cm}}$ |
| :--- | :---: | :---: | :---: |
| 2018 | 29.5 | 24 | 30.2 |
| 2019 | 30.1 | 24 | 30.2 |
| 2020 | 26.7 | 24 | 30.2 |
| 2021 | 26.3 | 22 | 30.2 |

### 26.6 Management plans

The EU has proposed a multiannual management plan for the Western Waters (EU, 2018). However, this stock was excluded from the final version (EU 2019, approved on 05/03/2019 Meeting n 03676 - https://www.consilium.europa.eu/en/meetings/env/2019/03/05/). Therefore, there is no management plan for Celtic Sea plaice.

### 26.7 Uncertainties in assessment and forecast

In spite of the COVID-19 pandemic, the sampling levels of landed catch and discards as seen in contribution of sampled and unsampled landings and discards to final assessment (Figure 26.3) in 2021 was sufficient to provide reliable information on the length- and age structure of the stock.

From 2003 onwards, discard sampling for Ireland, Belgium, France and the UK(E\&W) has been improved under the Data Collection Regulation. Discarding remained too high (exceeding landings) in this fishery until 2019, thereby compromising the effectiveness of quota management on landings. In 2019-2021 landings first time exceeded discards. It is difficult to predict fishing fleet behaviour, as it is a commercial species of a low value taken mostly as a bycatch to fishery for sole, and to lesser extent, to Nephrops.

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Table 26.1. Plaice in divisions 7.f-g. Nominal landings ( $\mathbf{t}$ ) as reported to ICES, and total landings as used by ICES WGCSE.

|  | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 214 | 196 | 171 | 372 | 365 | 341 | 314 | 283 | 357 | 665 |
| UK (Engl. \& Wales) | 150 | 152 | 176 | 227 | 251 | 196 | 279 | 366 | 466 | 529 |
| France | 365 | 527 | 467 | 706 | 697 | 568 | 532 | 558 | 493 | 878 |
| Ireland | 28 | 0 | 49 | 61 | 64 | 198 | 48 | 72 | 91 | 302 |
| $N$. Ireland |  |  |  |  |  |  |  |  |  |  |
| Netherlands |  |  |  |  |  |  |  |  |  | 9 |
| Scotland | 0 | 0 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 1 |
| Total reported | 757 | 875 | 863 | 1373 | 1377 | 1303 | 1173 | 1279 | 1407 | 2384 |
| Discards | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
| Unallocated | 0 | 0 | 0 | 0 | 0 | 0 | -27 | -69 | 345 | -693 |
| Landings used by WG | 757 | 875 | 863 | 1373 | 1377 | 1303 | 1146 | 1210 | 1752 | 1691 |
| Catch as used by WG | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| Belgium | 581 | 617 | 843 | 794 | 836 | 371 | 542 | 350 | 346 | 410 |
| UK (Engl. \& Wales) | 496 | 629 | 471 | 497 | 392 | 302 | 290 | 251 | 284 | 239 |
| France | 708 | 721 | 1089 | 767 | 444 | 504 | 373 | 298 | 254 | 246 |
| Ireland | 127 | 226 | 180 | 160 | 155 | 180 | 89 | 82 | 70 | 83 |
| $N$. Ireland |  | 1 |  |  |  |  |  |  |  |  |
| Scotland |  |  |  | 1 |  | 5 | 9 | 1 | 2 |  |
| Total reported | 1912 | 2194 | 2583 | 2219 | 1827 | 1362 | 1303 | 982 | 956 | 978 |
| Discards | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
| Unallocated | -11 | -78 | -432 | -137 | -326 | -174 | -189 | 88 | 72 | -26 |
| Landings used by WG | 1901 | 2116 | 2151 | 2082 | 1501 | 1188 | 1114 | 1070 | 1028 | 952 |
| Catch as used by WG | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| Belgium | 594 | 540 | 371 | 224 | 241 | 248 | 221 | 212 | 168 | 172 |
| UK (Engl. \& Wales) | 258 | 176 | 170 | 134 | 136 | 105 | 127 | 87 | 55 | 88 |
| France | 329 | 298 |  | 287 | 262 | 186 | 165 | 145 | 132 | 106 |


| Ireland | 78 | 135 | 115 | 76 | 45 | 79 | 51 | 45 | 44 | 48 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total reported | 1259 | 1149 | 656 | 721 | 684 | 618 | 564 | 489 | 399 | 414 |
| Discards | N/A | N/A | N/A | N/A | N/A | N/A | N/A | 274 | 321 | 453 |
| Unallocated | -42 | -82 | 312 | -3 | 30 | 24 | 30 | 21 | -13 | -10 |
| Landings used by WG | 1217 | 1067 | 968 | 718 | 714 | 642 | 594 | 510 | 386 | 404 |
| Catch as used by WG | N/A | N/A | N/A | N/A | N/A | N/A | N/A | 784 | 707 | 857 |
|  | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| Belgium | 194 | 188 | 216 | 188 | 210 | 203 | 185 | 182 | 185 | 243 |
| UK | 61 | 63 | 55 | 54 | 45 | 45 | 41 | 25 | 25 | 27 |
| France | 104 | 70 | NA | 136 | 98 | 125 | 106 | 155 | 111 | 108 |
| Ireland | 58 | 64 | 63 | 63 | 67 | 76 | 80 | 49 | 59 | 52 |
| Total reported | 417 | 385 | NA | 442 | 420 | 450 | 412 | 411 | 381 | 430 |
| Discards | 1288 | 583 | 608 | 670 | 1107 | 1123 | 1274 | 772 | 778 | 571 |
| Unallocated | -7 | 52 | -1 | -9 | 7 | -8 | -2 | -1 | 0 | 0 |
| Landings used by WG | 410 | 437 | 481 | 442 | 427 | 442 | 414 | 410 | 381 | 431 |
| Catch as used by WG | 1698 | 1020 | 1089 | 1112 | 1534 | 1565 | 1688 | 1183 | 1159 | 1002 |
|  | 2017 | 2018 | 2019 | 2020 | 2021 |  |  |  |  |  |
| Belgium | 179 | 204 | 263 | 332 | 263 |  |  |  |  |  |
| UK | 38 | 40 | 45 | 47 | 48 |  |  |  |  |  |
| France | 108 | 127 | 84 | 47 | 51 |  |  |  |  |  |
| Spain |  |  | 1 | <1 | 0.5 |  |  |  |  |  |
| Netherlands |  |  | <1 | <1 | <1 |  |  |  |  |  |
| Ireland | 63 | 51 | C* | 110 | 105 |  |  |  |  |  |
| Total reported | 388 | 422 | 394 C* | 536 | 468 |  |  |  |  |  |
| Discards | 895 | 508 | 189 | 357 | 378 |  |  |  |  |  |
| Unallocated | 1 | 0 | -2 | 0 | 2 |  |  |  |  |  |
| Landings used by WG | 389 | 422 | 642 | 536 | 470 |  |  |  |  |  |
| Catch as used by WG | 1284 | 930 | 831 | 893 | 468 |  |  |  |  |  |

C Incomplete / missing due to part of the data being unavailable under national GDPR clauses.

## Table 26.2. Plaice in divisions 7.f-g: lpue and cpue for UK(E\&W) fleets.



|  | LANDINGS PER UNIT OF EFFORT (LPUE) kg\day |  |  |  |  |  | EFFORT, fishing days |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | VIIf |  | VIIg EAST |  | VIIg WEST |  | VIIf |  | VIIg EAST |  | VIIg WEST |  |
|  | Ipue |  | Ipue |  | Ipue |  | Effort |  | Effort |  | Effort |  |
| YEAR | TRAWL | BEAM <br> TRAWL | TRAWL | BEAM | TRAWL | BEAM | TRAWL | BEAM | TRAWL | BEAM | TRAWL | BEAM |
|  |  |  |  | TRAWL |  | TRAWL | (Days fished) | (Days fished) | (Days fished) | (Days fished) | (Days fished) | (Days fished) |
| 1997 | 34.61 | 31.01 | 21.37 | 33.42 | 17.47 | 7.5 | 577 | 2661 | 370 | 770 | 122 | 146 |
| 1998 | 21.86 | 26.07 | 15.53 | 15.33 | 5.12 | 12.65 | 517 | 2846 | 385 | 591 | 94 | 159 |
| 1999 | 35.6 | 26.62 | 20.65 | 12 | 5.14 | 11.96 | 395 | 3058 | 176 | 1461 | 235 | 312 |
| 2000 | 32.09 | 16.1 | 40.58 | 11.64 | 3.35 | 10.1 | 284 | 3133 | 187 | 1007 | 160 | 200 |
| 2001 | 34.02 | 16.69 | 32.3 | 15.26 | 4.66 | 11.04 | 309 | 3172 | 187 | 1155 | 179 | 91 |
| 2002 | 19.78 | 15.64 | 48.8 | 20.81 | 7.43 | 4.81 | 416 | 2652 | 123 | 463 | 170 | 60 |
| 2003 | 23.45 | 18.24 | 8.19 | 20.78 | 4.48 | 1.49 | 696 | 2669 | 51 | 772 | 124 | 158 |
| 2004 | 18.77 | 15.54 | 8.66 | 7.81 | 3.09 | 3.39 | 641 | 2503 | 198 | 923 | 125 | 178 |
| 2005 | 11.2 | 11 | 2.14 | 8.25 | 0.25 | 1.33 | 876 | 1968 | 21 | 618 | 154 | 116 |
| 2006 | 21.21 | 12.77 | 5.91 | 15.19 | 0.64 | 0.58 | 924 | 1330 | 23 | 630 | 233 | 70 |
| 2007 | 14.79 | 17.93 | 20.42 | 10.58 | 1.71 | 5.9 | 798 | 1407 | 31 | 518 | 219 | 12 |
| 2008 | 18.01 | 21.2 | 21.1 | 10.22 | 0.08 | 1.72 | 711 | 1202 | 109 | 290 | 229 | 5 |
| 2009 | 14.4 | 15.66 | 11.58 | 14.77 | 1.63 | 0.76 | 656 | 1105 | 244 | 266 | 296 | 48 |
| 2010 | 14.09 | 27.93 | 12.88 | 11.82 | 0.31 | 1.06 | 565 | 1162 | 84 | 327 | 469 | 78 |
| 2011 | 11.11 | 32.98 | 5.43 | 17.11 | 2.09 | 0.76 | 525 | 868 | 8 | 180 | 353 | 111 |
| 2012 | 10.96 | 17.7 | 3.11 | 9.38 | 0.67 | 0.51 | 543 | 1408 | 138 | 275 | 487 | 102 |


|  | LANDINGS PER UNIT OF EFFORT (LPUE) kg\day |  |  |  |  |  | EFFORT, fishing days |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | VIIf |  | VIIg EAST |  | VIIg WEST |  | VIIf |  | VIIg EAST |  | VIIg WEST |  |
|  | Ipue |  | Ipue |  | Ipue |  | Effort |  | Effort |  | Effort |  |
| YEAR | TRAWL | BEAM | TRAWL | BEAM | TRAWL | BEAM | TRAWL |  | TRAWL | BEAM | TRAWL | BEAM |
|  | TRAWL |  | TRAWL |  | TRAWL |  | (Days fished) | (Days fished) | (Days fished) | (Days fished) | (Days fished) | (Days fished) |
| 2013 | 6.4 | 12.29 | 0.89 | 8.18 | 0.44 | 0.61 | 280 | 1611 | 72 | 265 | 37 | 77 |
| 2014 | 5.76 | 15.52 | 7.43 | 10.61 | 0.08 | 2.5 | 156 | 959 | 10 | 131 | 176 | 24 |
| 2015 | 18.82 | 11.87 | 37.87 | 14.58 | 0 | 3.65 | 79 | 726 | 3 | 245 | 165 | 56 |
| 2016 | 0 | 14.91 | 0 | 9.57 | 0.07 | 0.05 | 136 | 915 | 0 | 396 | 329 | 34 |
| 2017 | 24.98 | 18.57 | 2.97 | 10.28 | 0.05 | 2.47 | 93 | 986 | 95 | 514 | 193 | 74 |
| 2018 | 11.1 | 19.53 | 27.1 | 7.77 | 0.93 | 10.7 | 127 | 1071 | 71 | 440 | 210 | 15 |
| 2019 | 19.62 | 32.97 | 11.95 | 26.73 | 1.11 | 8.58 | 169 | 981 | 34 | 255 | 277 | 8 |
| 2020 | 8.4 | 34.41 | 2.31 | 26.41 | 0 | 0.17 | 100 | 1012 | 10 | 346 | 40 | 99 |
| 2021 | 6.36 | 23.68 | 35.1 | 17.14 | 17.35 | 0.16 | 155 | 1260 | 22 | 540 | 28 | 102 |

Table 26.3. Plaice in divisions 7.f-g: lpue and effort for Belgian fleets in 7.f-g.

| BELGIAN Beam Trawl 7fg |  |  |  |
| :---: | :---: | :---: | :---: |
| Year | Landings ( t ) | Effort (000 hr) | Ipue (kg/h) |
| 1996 | 356.89 | 53.27 | 6.70 |
| 1997 | 474.71 | 57.36 | 8.28 |
| 1998 | 443.38 | 57.79 | 7.67 |
| 1999 | 410.22 | 55.11 | 7.44 |
| 2000 | 230.63 | 51.34 | 4.49 |
| 2001 | 274.84 | 54.90 | 5.01 |
| 2002 | 259.80 | 49.60 | 5.24 |
| 2003 | 215.95 | 62.73 | 3.44 |
| 2004 | 207.27 | 78.73 | 2.63 |
| 2005 | 153.73 | 64.50 | 2.38 |
| 2006 | 134.44 | 50.28 | 2.67 |
| 2007 | 139.39 | 45.72 | 3.05 |
| 2008 | 106.29 | 26.71 | 3.70 |
| 2009 | 140.76 | 30.84 | 4.56 |
| 2010 | 127.15 | 32.74 | 3.88 |
| 2011 | 159.03 | 41.41 | 3.84 |
| 2012 | 165.73 | 46.25 | 3.58 |
| 2013 | 155.973 | 45.159 | 3.454 |
| 2014 | 155.317 | 31.271 | 4.967 |
| 2015 | 165.17 | 31.792 | 5.195 |
| 2016 | 212.01 | 32.34 | 6.556 |
| 2017 | 169.03 | 33.35 | 5.07 |
| 2018 | 186.861 | 31.48 | 5.94 |
| 2019 | 226.443 | 32.033 | 7.131 |
| 2020 | 293.355 | 41.699 | 7.035 |
| 2021 | 244.64 | 36.182 | 6.761 |

Table 26.4. Plaice in Divisions 7.f-g: lpue and effort for Irish otter trawl, beam and seine fleets in 7.g.

|  | IR-OTB-7G |  |  | IR-SCC-7G |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings (t) | Effort (000 hr) | Ipue (kg/h) | Landings (t) | Effort (000 hr) | Ipue (kg/h) |
| 1995 | 94.23 | 63.56 | 1.48 | 9.55 | 6.43 | 1.49 |
| 1996 | 133.66 | 60.04 | 2.23 | 14.20 | 9.73 | 1.46 |
| 1997 | 119.84 | 65.10 | 1.84 | 38.79 | 16.13 | 2.40 |
| 1998 | 96.72 | 72.30 | 1.34 | 21.38 | 14.94 | 1.43 |
| 1999 | 60.05 | 51.66 | 1.16 | 10.40 | 8.01 | 1.30 |
| 2000 | 26.78 | 60.60 | 0.47 | 11.40 | 9.90 | 1.15 |
| 2001 | 23.82 | 69.43 | 0.34 | 10.93 | 16.33 | 0.67 |
| 2002 | 42.30 | 77.69 | 0.54 | 16.42 | 20.86 | 0.79 |
| 2003 | 26.35 | 86.79 | 0.30 | 13.80 | 20.91 | 0.66 |
| 2004 | 26.62 | 96.99 | 0.27 | 5.04 | 19.38 | 0.26 |
| 2005 | 22.78 | 124.40 | 0.18 | 6.47 | 14.81 | 0.44 |
| 2006 | 25.17 | 119.23 | 0.21 | 5.10 | 14.79 | 0.34 |
| 2007 | 30.99 | 136.52 | 0.23 | 4.76 | 15.82 | 0.30 |
| 2008 | 39.17 | 125.81 | 0.31 | 8.38 | 11.65 | 0.72 |
| 2009 | 43.81 | 137.11 | 0.32 | 7.98 | 8.19 | 0.98 |
| 2010 | 44.29 | 140.65 | 0.31 | 10.71 | 9.69 | 1.11 |
| 2011 | 44.68 | 120.33 | 0.37 | 11.12 | 11.01 | 1.01 |
| 2012 | 43.21 | 121.08 | 0.35 | 18.41 | 14.15 | 1.30 |
| 2013 | 31.91 | 118.13 | 0.28 | 11.10 | 12.06 | 0.84 |
| 2014 | 26.00 | 127.40 | 0.22 | 7.60 | 12.00 | 0.61 |
| 2015 | 33.34 | 132.69 | 0.25 | 8.36 | 9.28 | 0.90 |
| 2016 | 34.80 | 148.17 | 0.23 | 9.37 | 10.44 | 0.90 |
| 2017 | 40.86 | 135.98 | 0.30 | 10.49 | 9.75 | 1.08 |
| 2018 | 33.64 | 105.81 | 0.32 | 8.13 | 9.69 | 0.84 |
| 2019 | 33.89 | 103.89 | 0.33 | 16.40 | 14.26 | 1.15 |
| 2020 | 54.63 | 89.91 | 0.61 | 17.45 | 13.59 | 1.28 |
| 2021 | 52.04 | 83.90 | 0.62 | 25.13 | 14.80 | 1.70 |


| IR-TBB-7G |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings (t) | Effort (000 hr) | Ipue (kg/h) | Year | Landings (t) | Effort (000 hr) | Ipue (kg/h) |
| 1995 | 37.92 | 20.78 | 1.83 | 2009 | 6.96 | 37.96 | 0.18 |
| 1996 | 53.02 | 26.76 | 1.98 | 2010 | 6.56 | 40.22 | 0.16 |
| 1997 | 94.59 | 26.25 | 3.35 | 2011 | 6.71 | 35.33 | 0.19 |
| 1998 | 122.13 | 35.25 | 3.46 | 2012 | 33.63 | 40.33 | 0.83 |
| 1999 | 25.80 | 40.87 | 0.63 | 2013 | 32.32 | 38.48 | 0.84 |
| 2000 | 12.62 | 37.03 | 0.34 | 2014 | 12.50 | 37.80 | 0.33 |
| 2001 | 4.80 | 39.71 | 0.12 | 2015 | 12.10 | 37.79 | 0.32 |
| 2002 | 7.08 | 31.62 | 0.22 | 2016 | 9.83 | 39.55 | 0.25 |
| 2003 | 9.37 | 49.26 | 0.19 | 2017 | 12.39 | 35.21 | 0.35 |
| 2004 | 6.17 | 54.86 | 0.11 | 2018 | 9.62 | 37.42 | 0.26 |
| 2005 | 9.49 | 49.65 | 0.19 | 2019 | 20.32 | 34.08 | 0.60 |
| 2006 | 14.46 | 60.48 | 0.24 | 2020 | 43.20 | 29.14 | 1.48 |
| 2007 | 21.18 | 55.86 | 0.38 | 2021 | 49.07 | 31.57 | 1.55 |
| 2008 | 14.18 | 37.22 | 0.38 |  |  |  |  |

Table 26.5. Plaice in divisions 7.f-g. Landings numbers-at-age.


| 6 | 121 | 96 | 116 | 145 | 108 | 113 | 85 | 46 | 73 | 76 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | 60 | 82 | 45 | 53 | 62 | 65 | 34 | 27 | 24 | 50 |
| 8 | 20 | 39 | 27 | 22 | 23 | 24 | 26 | 15 | 14 | 12 |
| +gp | 82 | 56 | 69 | 32 | 28 | 28 | 24 | 17 | 16 | 15 |
| TOTALNUM | 3231 | 2773 | 2678 | 1881 | 1838 | 1789 | 1608 | 1504 | 1078 | 1229 |
| AGE\YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| 1 | 8 | 15 | 2 | 3 | 1 | 2 | 3 | 0 | 0 | 2 |
| 2 | 130 | 270 | 127 | 135 | 135 | 106 | 64 | 24 | 55 | 20 |
| 3 | 513 | 341 | 626 | 223 | 326 | 485 | 328 | 123 | 122 | 332 |
| 4 | 340 | 443 | 345 | 430 | 208 | 288 | 383 | 452 | 231 | 201 |
| 5 | 104 | 145 | 273 | 191 | 248 | 164 | 192 | 247 | 410 | 182 |
| 6 | 76 | 47 | 68 | 152 | 130 | 163 | 67 | 109 | 127 | 228 |
| 7 | 46 | 29 | 20 | 44 | 69 | 65 | 70 | 33 | 43 | 94 |
| 8 | 26 | 11 | 10 | 8 | 28 | 33 | 29 | 36 | 17 | 42 |
| +gp | 13 | 15 | 12 | 8 | 17 | 23 | 31 | 30 | 26 | 37 |
| TOTALNUM | 1257 | 1315 | 1485 | 1187 | 1161 | 1329 | 1167 | 1054 | 1052 | 1138 |
| AGE\YEAR | 2017 | 2018 | 2019 | 2020 | 2021 |  |  |  |  |  |
| 1 | 0 | 3 | 4 | 1 | 0 |  |  |  |  |  |
| 2 | 33 | 32 | 28 | 98 | 56 |  |  |  |  |  |
| 3 | 57 | 143 | 85 | 155 | 339 |  |  |  |  |  |
| 4 | 380 | 122 | 248 | 190 | 328 |  |  |  |  |  |
| 5 | 167 | 393 | 187 | 256 | 147 |  |  |  |  |  |
| 6 | 112 | 160 | 336 | 229 | 179 |  |  |  |  |  |
| 7 | 145 | 92 | 215 | 337 | 93 |  |  |  |  |  |
| 8 | 56 | 89 | 63 | 163 | 134 |  |  |  |  |  |
| +gp | 35 | 62 | 83 | 145 | 136 |  |  |  |  |  |
| TOTALNUM | 985 | 1096 | 1249 | 1574 | 1412 |  |  |  |  |  |

Table 26.6. Plaice in divisions 7.f-g. Landings weights-at-age.

| Landings weights-at-age (kg) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 0.078 | 0.194 | 0.076 | 0.118 | 0.185 | 0.151 | 0.178 | 0.276 | 0.135 | 0.000 |
| 2 | 0.205 | 0.258 | 0.203 | 0.238 | 0.255 | 0.245 | 0.274 | 0.324 | 0.251 | 0.160 |
| 3 | 0.323 | 0.323 | 0.325 | 0.354 | 0.330 | 0.339 | 0.369 | 0.384 | 0.363 | 0.301 |
| 4 | 0.430 | 0.389 | 0.440 | 0.467 | 0.412 | 0.433 | 0.464 | 0.455 | 0.470 | 0.434 |
| 5 | 0.528 | 0.457 | 0.550 | 0.576 | 0.500 | 0.526 | 0.559 | 0.538 | 0.572 | 0.559 |
| 6 | 0.615 | 0.525 | 0.652 | 0.682 | 0.595 | 0.620 | 0.654 | 0.633 | 0.670 | 0.677 |
| 7 | 0.693 | 0.595 | 0.749 | 0.784 | 0.695 | 0.714 | 0.749 | 0.739 | 0.763 | 0.787 |
| 8 | 0.760 | 0.666 | 0.839 | 0.882 | 0.802 | 0.808 | 0.844 | 0.857 | 0.851 | 0.889 |
| +gp | 0.8762 | 0.8435 | 1.0653 | 1.1812 | 1.1824 | 1.0948 | 1.1579 | 1.2661 | 1.0036 | 1.1033 |
| SOPCOFAC | 1.0052 | 1.0262 | 1.0225 | 1.0135 | 1.0042 | 1.0125 | 0.9995 | 1.0000 | 1.0047 | 0.9997 |
| AGE\YEAR | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 0.129 | 0.260 | 0.102 | 0.240 | 0.200 | 0.148 | 0.171 | 0.236 | 0.219 | 0.000 |
| 2 | 0.208 | 0.288 | 0.176 | 0.270 | 0.260 | 0.257 | 0.263 | 0.296 | 0.254 | 0.247 |
| 3 | 0.288 | 0.325 | 0.255 | 0.309 | 0.327 | 0.362 | 0.314 | 0.308 | 0.304 | 0.295 |
| 4 | 0.368 | 0.370 | 0.337 | 0.358 | 0.400 | 0.464 | 0.405 | 0.397 | 0.364 | 0.349 |
| 5 | 0.449 | 0.423 | 0.423 | 0.416 | 0.481 | 0.563 | 0.500 | 0.455 | 0.485 | 0.512 |
| 6 | 0.530 | 0.484 | 0.514 | 0.483 | 0.567 | 0.658 | 0.598 | 0.598 | 0.603 | 0.553 |
| 7 | 0.612 | 0.554 | 0.608 | 0.560 | 0.661 | 0.750 | 0.643 | 0.801 | 0.714 | 0.523 |
| 8 | 0.694 | 0.633 | 0.706 | 0.646 | 0.761 | 0.839 | 0.728 | 0.728 | 0.752 | 0.947 |
| +gp | 0.8632 | 0.8887 | 0.9932 | 0.9097 | 1.0465 | 1.0399 | 0.9886 | 0.9585 | 1.0655 | 1.0667 |
| SOPCOFAC | 1.0034 | 1.0024 | 1.0006 | 1.0009 | 1.0113 | 1.0022 | 0.9997 | 1.0001 | 1.0004 | 0.9998 |
| AGE\YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| 1 | 0.249 | 0.213 | 0.213 | 0.245 | 0.268 | 0.246 | 0.205 | 0.221 | 0.237 | 0.238 |
| 2 | 0.291 | 0.256 | 0.268 | 0.260 | 0.305 | 0.284 | 0.295 | 0.258 | 0.260 | 0.246 |
| 3 | 0.304 | 0.317 | 0.278 | 0.302 | 0.340 | 0.281 | 0.321 | 0.287 | 0.295 | 0.291 |
| 4 | 0.357 | 0.380 | 0.332 | 0.370 | 0.398 | 0.343 | 0.353 | 0.330 | 0.356 | 0.339 |
| 5 | 0.466 | 0.463 | 0.440 | 0.479 | 0.466 | 0.433 | 0.439 | 0.382 | 0.425 | 0.385 |


| 6 | 0.663 | 0.604 | 0.538 | 0.539 | 0.556 | 0.484 | 0.502 | 0.514 | 0.525 | 0.513 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | 0.745 | 0.661 | 0.618 | 0.672 | 0.675 | 0.541 | 0.651 | 0.649 | 0.631 | 0.549 |
| 8 | 0.877 | 0.690 | 0.839 | 0.875 | 0.695 | 0.859 | 0.681 | 0.750 | 0.714 | 0.638 |
| +gp | 1.1007 | 1.1886 | 1.1906 | 1.2018 | 1.0905 | 1.1262 | 1.0389 | 0.9919 | 1.0163 | 0.8369 |
| SOPCOFAC | 1.0002 | 1.0009 | 1.0000 | 1.0007 | 1.0007 | 1.0004 | 0.9994 | 1.0007 | 1.0011 | 1.0008 |
| AGE\YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| 1 | 0.278 | 0.260 | 0.279 | 0.233 | 0.228 | 0.235 | 0.273 | 0.156 | 0.15 | 0.211 |
| 2 | 0.271 | 0.273 | 0.267 | 0.292 | 0.242 | 0.246 | 0.285 | 0.28 | 0.24 | 0.253 |
| 3 | 0.277 | 0.298 | 0.275 | 0.331 | 0.283 | 0.280 | 0.286 | 0.312 | 0.275 | 0.278 |
| 4 | 0.303 | 0.329 | 0.329 | 0.328 | 0.335 | 0.307 | 0.320 | 0.346 | 0.3 | 0.318 |
| 5 | 0.389 | 0.386 | 0.376 | 0.376 | 0.378 | 0.345 | 0.370 | 0.386 | 0.365 | 0.365 |
| 6 | 0.457 | 0.433 | 0.469 | 0.458 | 0.465 | 0.418 | 0.465 | 0.504 | 0.467 | 0.416 |
| 7 | 0.537 | 0.511 | 0.499 | 0.598 | 0.600 | 0.498 | 0.517 | 0.473 | 0.514 | 0.510 |
| 8 | 0.547 | 0.719 | 0.605 | 0.469 | 0.690 | 0.570 | 0.602 | 0.599 | 0.609 | 0.567 |
| +gp | 0.9862 | 0.9042 | 0.7197 | 1.0433 | 1.1810 | 0.6750 | 0.6550 | 0.735 | 0.946 | 1.003 |
| SOPCOFAC | 1.0005 | 1.0001 | 0.9993 | 1.0002 | 1.0000 | 1.0001 | 0.9994 | 1.001 | 1.002 | 1.005 |
| AGE\YEAR | 2017 | 2018 | 2019 | 2020 | 2021 |  |  |  |  |  |
| 1 | 0.231 | 0.198 | 0.206 | 0.185 |  |  |  |  |  |  |
| 2 | 0.279 | 0.229 | 0.231 | 0.225 | 0.221 |  |  |  |  |  |
| 3 | 0.289 | 0.262 | 0.277 | 0.245 | 0.256 |  |  |  |  |  |
| 4 | 0.325 | 0.297 | 0.306 | 0.275 | 0.272 |  |  |  |  |  |
| 5 | 0.370 | 0.326 | 0.337 | 0.31 | 0.311 |  |  |  |  |  |
| 6 | 0.426 | 0.407 | 0.377 | 0.318 | 0.346 |  |  |  |  |  |
| 7 | 0.460 | 0.468 | 0.376 | 0.358 | 0.403 |  |  |  |  |  |
| 8 | 0.590 | 0.515 | 0.513 | 0.415 | 0.442 |  |  |  |  |  |
| +gp | 0.7620 | 0.739 | 0.809 | 0.594 | 0.551 |  |  |  |  |  |
| SOPCOFAC | 1.0400 | 0.978 | 1.03 | 1.005 | 0.999 |  |  |  |  |  |

Table 26.7. Plaice in divisions 7.f-g. Discards numbers-at-age.

| Discard numbers-at-age Numbers $10^{3}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 1 | 0 | 0 | 0 | 455 | 572 | 542 | 1829 | 73 | 671 | 385 | 960 | 142 | 614 | 83 |
| 2 | 0 | 0 | 0 | 360 | 1211 | 2584 | 3331 | 3595 | 985 | 2719 | 2656 | 2496 | 1283 | 987 |
| 3 | 0 | 0 | 0 | 641 | 441 | 750 | 3408 | 632 | 2041 | 1017 | 1429 | 1950 | 3581 | 1672 |
| 4 | 0 | 0 | 0 | 171 | 118 | 74 | 814 | 393 | 761 | 550 | 1019 | 502 | 1004 | 3195 |
| 5 | 0 | 0 | 0 | 68 | 41 | 47 | 81 | 69 | 399 | 345 | 501 | 179 | 231 | 454 |
| 6 | 0 | 0 | 0 | 3 | 12 | 12 | 32 | 4 | 44 | 54 | 45 | 163 | 32 | 173 |
| 7 | 0 | 0 | 0 | 4 | 4 | 1 | 11 | 1 | 4 | 8 | 99 | 58 | 44 | 77 |
| 8 | 0 | 0 | 0 | 1 | 22 | 1 | 9 | 1 | 5 | 0 | 56 | 25 | 11 | 27 |
| +gp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 8 | 36 |
| TOTALNUM | 0 | 0 | 0 | 1703 | 2421 | 4011 | 9515 | 4768 | 4910 | 5078 | 6765 | 5523 | 6808 | 6704 |
| TONSLAND | 0 | 0 | 0 | 274 | 321 | 453 | 1288 | 583 | 608 | 670 | 1107 | 852 | 1260 | 1158 |
| SOPCOFAC | 0 | 0 | 0 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |


| Discard numbers-at-age Numbers 10 |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| AGE\YEAR | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| 0 | - | - | 0 | 0 | 0 | 0 | 0 |
| 1 | 38 | 29 | 169 | 3 | 29 | 13 | 0 |
| 2 | 1527 | 224 | 739 | 92 | 260 | 448 | 273 |
| 3 | 1253 | 1610 | 1078 | 587 | 157 | 673 | 1114 |
| 4 | 753 | 615 | 1257 | 444 | 328 | 388 | 643 |
| 5 | 1106 | 229 | 478 | 668 | 149 | 394 | 172 |
| 7 | 303 | 209 | 312 | 346 | 111 | 169 | 165 |
| 8 | 54 | 34 | 147 | 307 | 38 | 148 | 154 |
| TOT | 33 | 15 | 32 | 11 | 0 | 32 | 51 |
| TOTALNUM | 80 | 9 | 13 | 103 | 8 | 26 | 18 |
| TONSLAND | 5145 | 2974 | 4225 | 2561 | 1080 | 2291 | 2590 |
| SOPCOF \% | 870 | 591 | 895 | 508 | 189 | 536 | 378 |

Table 26.8. Plaice in divisions 7.f-g. Discards weights-at-age.

| Discard weights-at-age (kg) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 1 | 0 | 0 | 0 | 0.123 | 0.095 | 0.064 | 0.088 | 0.092 | 0.088 | 0.085 | 0.118 | 0.104 | 0.097 | 0.040 |
| 2 | 0 | 0 | 0 | 0.152 | 0.127 | 0.107 | 0.126 | 0.11 | 0.127 | 0.125 | 0.148 | 0.124 | 0.129 | 0.112 |
| 3 | 0 | 0 | 0 | 0.177 | 0.154 | 0.154 | 0.159 | 0.154 | 0.127 | 0.143 | 0.173 | 0.167 | 0.180 | 0.160 |
| 4 | 0 | 0 | 0 | 0.194 | 0.188 | 0.176 | 0.163 | 0.172 | 0.127 | 0.149 | 0.168 | 0.192 | 0.233 | 0.181 |
| 5 | 0 | 0 | 0 | 0.212 | 0.202 | 0.201 | 0.204 | 0.211 | 0.143 | 0.163 | 0.225 | 0.239 | 0.277 | 0.214 |
| 6 | 0 | 0 | 0 | 0.337 | 0.344 | 0.242 | 0.249 | 0.282 | 0.194 | 0.189 | 0.304 | 0.247 | 0.459 | 0.227 |
| 7 | 0 | 0 | 0 | 0.23 | 0.403 | 0.395 | 0.368 | 0.365 | 0.2 | 0.445 | 0.339 | 0.238 | 0.380 | 0.300 |
| 8 | 0 | 0 | 0 | 0.455 | 0.419 | 0.349 | 0.425 | 0.283 | 0.257 | 0.523 | 0.389 | 0.337 | 0.312 | 0.470 |
| +gp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.713 |


| Discard weights-at-age (kg) |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| AGE\YEAR | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| 0 |  |  | 0.058 |  |  |  |  |
| 1 | 0.12 | 0.148 | 0.14 | 0.105 | 0.084 | 0.095 | 0.081 |
| 2 | 0.124 | 0.153 | 0.147 | 0.126 | 0.118 | 0.127 | 0.127 |
| 3 | 0.143 | 0.177 | 0.186 | 0.150 | 0.169 | 0.143 | 0.139 |
| 4 | 0.171 | 0.205 | 0.225 | 0.188 | 0.196 | 0.161 | 0.146 |
| 5 | 0.219 | 0.261 | 0.258 | 0.182 | 0.180 | 0.172 | 0.158 |
| 6 | 0.315 | 0.288 | 0.324 | 0.207 | 0.183 | 0.18 | 0.183 |
| 7 | 0.208 | 0.341 | 0.271 | 0.324 | 0.159 | 0.187 | 0.16 |
| 8 | 0.204 | 0.416 | 0.29 | 0.350 | 0.258 | 0.298 | 0.163 |
|  | 0.529 | 0.462 | 0.442 | 0.873 | 0.182 | 0.196 | 0.177 |

Table 26.9. Plaice in divisions 7.f-g. Stock weights-at-age.

| Stock weights-at-age (kg) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 0.112 | 0.086 | 0.107 | 0.109 | 0.082 | 0.096 | 0.103 | 0.256 | 0.075 | 0.000 |
| 2 | 0.216 | 0.170 | 0.212 | 0.217 | 0.167 | 0.192 | 0.206 | 0.298 | 0.193 | 0.087 |
| 3 | 0.315 | 0.252 | 0.313 | 0.322 | 0.257 | 0.288 | 0.307 | 0.352 | 0.307 | 0.232 |
| 4 | 0.406 | 0.334 | 0.412 | 0.426 | 0.350 | 0.383 | 0.408 | 0.418 | 0.417 | 0.369 |
| 5 | 0.492 | 0.414 | 0.507 | 0.528 | 0.447 | 0.479 | 0.507 | 0.495 | 0.521 | 0.498 |
| 6 | 0.570 | 0.493 | 0.599 | 0.628 | 0.548 | 0.574 | 0.606 | 0.584 | 0.621 | 0.619 |
| 7 | 0.642 | 0.570 | 0.689 | 0.727 | 0.653 | 0.668 | 0.704 | 0.685 | 0.717 | 0.733 |
| 8 | 0.707 | 0.646 | 0.775 | 0.823 | 0.762 | 0.763 | 0.801 | 0.797 | 0.808 | 0.839 |
| +gp | 0.839 | 0.822 | 1.015 | 1.132 | 1.129 | 1.049 | 1.114 | 1.190 | 0.965 | 1.064 |


| AGE\YEAR | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.089 | 0.249 | 0.066 | 0.228 | 0.173 | 0.092 | 0.171 | 0.236 | 0.219 | 0.000 |
| 2 | 0.168 | 0.273 | 0.139 | 0.254 | 0.229 | 0.203 | 0.263 | 0.296 | 0.254 | 0.247 |
| 3 | 0.248 | 0.305 | 0.215 | 0.288 | 0.293 | 0.310 | 0.314 | 0.308 | 0.304 | 0.295 |
| 4 | 0.328 | 0.346 | 0.295 | 0.332 | 0.363 | 0.414 | 0.405 | 0.397 | 0.364 | 0.349 |
| 5 | 0.408 | 0.395 | 0.380 | 0.386 | 0.440 | 0.514 | 0.500 | 0.455 | 0.485 | 0.512 |
| 6 | 0.489 | 0.453 | 0.468 | 0.448 | 0.523 | 0.611 | 0.598 | 0.598 | 0.603 | 0.553 |
| 7 | 0.571 | 0.518 | 0.560 | 0.520 | 0.613 | 0.705 | 0.643 | 0.801 | 0.714 | 0.523 |
| 8 | 0.822 | 0.837 | 0.938 | 0.854 | 0.987 | 1.000 | 0.989 | 0.959 | 1.066 | 1.067 |
| $+g p$ | 0.657 | 0.602 | 0.710 | 0.795 | 0.728 | 0.728 | 0.752 | 0.947 |  |  |


| AGE\YEAR | 1997 | 1998 | 1999 | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.249 | 0.213 | 0.213 | 0.245 | 0.268 | 0.246 | 0.205 | 0.221 | 0.237 | 0.238 |
| 2 | 0.291 | 0.256 | 0.268 | 0.260 | 0.305 | 0.284 | 0.295 | 0.258 | 0.260 | 0.246 |
| 3 | 0.304 | 0.317 | 0.278 | 0.302 | 0.340 | 0.281 | 0.321 | 0.287 | 0.295 | 0.291 |
| 4 | 0.357 | 0.380 | 0.332 | 0.370 | 0.398 | 0.343 | 0.353 | 0.330 | 0.356 | 0.339 |
| 5 | 0.466 | 0.463 | 0.440 | 0.479 | 0.466 | 0.433 | 0.439 | 0.382 | 0.425 | 0.385 |
| 6 | 0.663 | 0.604 | 0.538 | 0.539 | 0.556 | 0.484 | 0.502 | 0.514 | 0.525 | 0.513 |


| 7 | 0.745 | 0.661 | 0.618 | 0.672 | 0.675 | 0.541 | 0.651 | 0.649 | 0.631 | 0.549 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 8 | 0.877 | 0.690 | 0.839 | 0.875 | 0.695 | 0.859 | 0.681 | 0.750 | 0.714 | 0.638 |
| +gp | 1.101 | 1.189 | 1.191 | 1.202 | 1.091 | 1.126 | 1.039 | 0.992 | 1.016 | 0.837 |
| AGE\YEAR | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| 1 | 0.278 | 0.260 | 0.279 | 0.233 | 0.228 | 0.106 | 0.098 | 0.095 | 0.129 | 0.153 |
| 2 | 0.271 | 0.273 | 0.267 | 0.292 | 0.242 | 0.129 | 0.136 | 0.116 | 0.128 | 0.161 |
| 3 | 0.277 | 0.298 | 0.275 | 0.331 | 0.283 | 0.190 | 0.188 | 0.171 | 0.155 | 0.194 |
| 4 | 0.303 | 0.329 | 0.329 | 0.328 | 0.335 | 0.234 | 0.257 | 0.202 | 0.202 | 0.233 |
| 5 | 0.389 | 0.386 | 0.376 | 0.376 | 0.378 | 0.290 | 0.319 | 0.275 | 0.259 | 0.307 |
| 6 | 0.457 | 0.433 | 0.469 | 0.458 | 0.465 | 0.332 | 0.463 | 0.334 | 0.36 | 0.355 |
| 7 | 0.537 | 0.511 | 0.499 | 0.598 | 0.600 | 0.375 | 0.465 | 0.353 | 0.343 | 0.465 |
| 8 | 0.547 | 0.719 | 0.605 | 0.469 | 0.690 | 0.470 | 0.525 | 0.543 | 0.339 | 0.527 |
| +gp | 0.986 | 0.904 | 0.720 | 1.043 | 1.181 | 0.549 | 0.654 | 0.594 | 0.563 | 0.998 |
| AGE\YEAR | 2017 | 2018 | 2019 | 2020 | 2021 |  |  |  |  |  |
| 0 | 0.058 |  |  |  |  |  |  |  |  |  |
| 1 | 0.14 | 0.150 | 0.099 | 0.102 | 0.081 |  |  |  |  |  |
| 2 | 0.153 | 0.152 | 0.129 | 0.144 | 0.143 |  |  |  |  |  |
| 3 | 0.191 | 0.172 | 0.207 | 0.162 | 0.166 |  |  |  |  |  |
| 4 | 0.248 | 0.212 | 0.243 | 0.198 | 0.189 |  |  |  |  |  |
| 5 | 0.286 | 0.235 | 0.267 | 0.227 | 0.228 |  |  |  |  |  |
| 6 | 0.350 | 0.270 | 0.328 | 0.26 | 0.268 |  |  |  |  |  |
| 7 | 0.365 | 0.357 | 0.344 | 0.306 | 0.251 |  |  |  |  |  |
| 8 | 0.482 | 0.498 | 0.512 | 0.396 | 0.365 |  |  |  |  |  |
| +gp | 0.675 | 0.838 | 0.588 | 0.555 | 0.507 |  |  |  |  |  |

Table 26.10. Plaice in divisions 7.f-g: Survey abundance indices (numbers per hour).

| IGFS (Numbers per hour) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 2021 |  |  |  |  |  |  |
| 1 | 1 | 0.79 | 0.92 |  |  |  |  |
| 1 | 7 |  |  |  |  |  |  |
| 1 | 0.0 | 3.2 | 6.0 | 2.7 | 0.6 | 0.2 | 0.1 |
| 1 | 0.1 | 0.4 | 1.9 | 3.1 | 1.2 | 0.8 | 0.1 |
| 1 | 2.8 | 4.4 | 5.9 | 1.3 | 0.7 | 0.2 | 0.2 |
| 1 | 0.2 | 6.0 | 4.6 | 1.2 | 1.0 | 0.6 | 0.7 |
| 1 | 0.1 | 2.6 | 8.5 | 3.5 | 1.1 | 0.5 | 0.4 |
| 1 | 0.4 | 6.0 | 5.6 | 3.8 | 1.0 | 0.4 | 0.2 |
| 1 | 12.5 | 11.7 | 32.3 | 14.6 | 5.9 | 1.2 | 0.9 |
| 1 | 10.1 | 37.9 | 13.2 | 20.8 | 8.6 | 3.7 | 1.0 |
| 1 | 10.8 | 49.5 | 30.2 | 8.4 | 9.1 | 3.6 | 4.6 |
| 1 | 14.6 | 40.5 | 36.8 | 11.3 | 2.1 | 2.0 | 2.9 |
| 1 | 1.5 | 16.1 | 37.3 | 19.7 | 7.2 | 1.9 | 6.2 |
| 1 | 0.4 | 7.9 | 14.3 | 13.6 | 6.1 | 3.4 | 2.2 |
| 1 | 0.8 | 37.8 | 26.2 | 13.0 | 15.2 | 3.0 | 5.0 |
| 1 | 1.1 | 13.8 | 33.6 | 13.9 | 9.2 | 9.0 | 4.2 |
| 1 | 0.8 | 11.5 | 12.8 | 13.0 | 10.8 | 3.7 | 4.6 |
| 1 | 0.1 | 5.5 | 9.8 | 6.6 | 7.9 | 3.2 | 3.2 |
| 1 | 1.6 | 7.2 | 5.8 | 13.1 | 5.8 | 5.1 | 4.0 |
| 1 | 0.8 | 8.1 | 7.0 | 5.1 | 5.9 | 3.8 | 6.6 |
| 1 | 0.6 | 27.7 | 30.5 | 17.4 | 5.6 | 4.4 | 5.1 |


| E+W BTS (numbers per 100 km towed) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 19952021 |  |  |  |  |  |
| 110.750 .85 |  |  |  |  |  |
| 15 |  |  |  |  |  |
| 1 | 239.590 | 90.480 | 17.230 | 2.960 | 6.840 |
| 1 | 223.690 | 288.110 | 30.780 | 0.990 | 2.620 |
| 1 | 225.370 | 102.140 | 34.540 | 4.250 | 1.770 |
| 1 | 237.200 | 126.220 | 46.990 | 8.920 | 2.000 |
| 1 | 152.590 | 79.620 | 29.030 | 19.670 | 7.000 |
| 1 | 339.630 | 63.170 | 31.250 | 6.560 | 5.500 |
| 1 | 211.440 | 156.140 | 15.810 | 8.740 | 4.230 |
| 1 | 136.740 | 175.120 | 80.450 | 5.930 | 6.130 |
| 1 | 98.370 | 80.480 | 60.950 | 21.830 | 2.720 |
| 1 | 258.510 | 33.410 | 27.080 | 13.420 | 2.190 |
| 1 | 192.500 | 75.220 | 20.870 | 8.060 | 10.930 |
| 1 | 85.780 | 101.970 | 34.160 | 9.570 | 1.790 |
| 1 | 150.400 | 92.250 | 47.260 | 15.110 | 1.670 |
| 1 | 140.690 | 217.040 | 46.790 | 15.700 | 4.820 |
| 1 | 161.810 | 55.960 | 78.580 | 21.450 | 10.890 |
| 1 | 331.760 | 88.540 | 26.410 | 39.940 | 6.680 |
| 1 | 362.260 | 300.140 | 55.040 | 21.860 | 21.370 |
| 1 | 142.130 | 430.790 | 100.570 | 22.360 | 9.020 |
| 1 | 329.790 | 139.060 | 185.390 | 46.850 | 5.770 |
| 1 | 371.760 | 202.300 | 64.650 | 105.700 | 23.800 |
| 1 | 26.360 | 454.080 | 162.340 | 52.370 | 76.660 |
| 1 | 12.520 | 163.100 | 268.260 | 102.300 | 27.500 |
| 1 | 11.490 | 104.1 | 137.39 | 121.110 | 91.87 |
| 1 | 4.15 | 45.26 | 90.2 | 58.1 | 75.08 |
| 1 | 114.94 | 138.97 | 38.18 | 15.37 | 11.19 |
| 1 | 7.17 | 113.19 | 139.61 | 42.71 | 11.95 |
| 1 | 37.87 | 44.4 | 76.06 | 42.6 | 7.11 |

Table 26.11. SAM-estimated reference points.

| reference point | value at low $\mathbf{m}$ | value at high $\mathbf{m}$ | rationale |
| :--- | :---: | :---: | :--- |
| MSY $_{\text {trigger }}$ | 1882 t | $2,885 \mathrm{t}$ | Lower 5th percentile of $\mathrm{B}_{\text {MSY; in tonnes }}$ |
| $\mathrm{F}_{\text {MSY }}$ | 0.147 | 0.468 | Stochastic simulations with segmented regression |
| $\mathrm{F}_{\text {MSYLower }}$ | 0.098 | 0.258 | Median lower point estimates of Stochastic simulations |
| $\mathrm{F}_{\text {MSYUpper }}$ | 0.228 | 0.803 | Median upper point estimates of Stochastic simulations |
| $\mathrm{B}_{\text {lim }}$ | 1344 t | 2061 t | Lowest observed SSB |
| $\mathrm{B}_{\text {pa }}$ | 1882 t | 2885 t | $\mathrm{B}_{\text {lim }}$ combined with the assessment error |
| $\mathrm{F}_{\text {lim }}$ | 1.17 | 1.51 | F with 50\% probability of SSB less than $\mathrm{B}_{\text {lim }}$ |
| $\mathrm{F}_{\text {pa }}$ | 0.84 | 1.08 | $\mathrm{~F}_{\text {lim }}$ combined with the assessment error |



Figure 26.1. Spatial distribution of plaice landings in the area 7 fg in 2009-2020 showing relative importance of the different ICES rectangles. Circle size is proportional to the percentage of the total annual landings taken from a particular rectangle.


Figure 26.2. Plaice in divisions 7.f-g: Landing and discards by different metiers when both landings and discards were sampled simultaneously. As the minimum conservation reference size is 270 mm , it is seen that no nearly no fish of commercial size was discarded.


Figure 26.3. Plaice in divisions 7.f-g: Contribution of sampled and unsampled landings and discards to final assessment catch numbers-at-age in 2021.


Figure 26.4. Plaice in divisions 7.f-g: Age composition of international landings (blue) and discards (red) from 2004 to 2021.


Figure 26.5. Plaice in divisions 7.f-g: Trends standardised cpues (kg/100 km towed) of surveys.

Vllfg plaice: UK(E\&W) LPUE


Figure 26.6. Plaice in divisions 7.f-g: Trends in CPUE by the UK fleets.


Figure 26.7. Plaice in divisions 7.f-g: Trends in CPUE by the EU fleets.


Figure 26.8. Plaice in divisions 7.f-g: Raised size structure of landings and discards in 2018-2021.


Figure 26.9. Plaice in divisions 7.f-g: Performance of the SAM model at the benchmark 2022.


Figure 26.10. Plaice in divisions 7.f-g: Performance of the SAM model with correlation of fishing mortalities between ages.


Figure 26.11. Plaice in divisions 7.f-g: Retrospective pattern in SAM with correlation of fishing mortalities between ages.


Figure 26.12. Ple 7 f\&g. Growth curves in 2000-2020 (data for 2021 are scarce).


Figure 26.13. Ple $7 \mathrm{f} \& \mathrm{~g}$. Position of mean catch length in respect to the length indicator $\mathrm{L}=\mathrm{m}$.

# 27 Plaice (Pleuronectes platessa) in divisions 7h-k (Celtic Sea South, southwest of Ireland) 

## Type of assessment in 2022

A trends-based assessment was conducted using a combined survey index produced using VAST biomass index of the VAST model to which an 'rfb' advice rule was applied. Stock status, FMSY proxy reference point, was estimated from LBI analysis, assuming $\mathrm{M} / \mathrm{K}=1.5$. $\mathrm{LF}=\mathrm{m}$ is based on $\mathrm{L}_{\mathrm{c}}$ (length at $50 \%$ of modal abundance), which varies each year.

## ICES advice applicable to 2023 and 2024

ICES advises that when the MSY approach is applied, catches should be no more than 132 tonnes in each of the years 2023 and 2024.

### 27.1 Impact of the COVID-19 pandemic

The sampling of plaice in $27.7 \mathrm{~h}-\mathrm{k}$ was effected by the implications of the COVID-19 pandemic in 2020 and 2021. Sampling levels were lower than previous years due to national restrictions on scientific sampling activities. There were insufficient samples to estimate discards based on 2021 data only, therefore a 5-year geometric mean was used (2017-2021).

Data for all scientific tuning indices were received for 2021.

### 27.2 General

### 27.2.1 Stock description and management units

The TAC specified for plaice in ICES Division $7 . \mathrm{h}-\mathrm{k}$ is consistent with the assessment area. Official national landings data as reported to ICES and the landings estimates as used by the Working Group are given in Table 27.1. Official landings in 2021 were 46 t , a $31 \%$ undershoot of the 2021 TAC ( 67 t ).

Plaice in 7.h-k is on the south-western margins of the species distribution. Landings of plaice are similar in ICES divisions 7 h and 7 j , but are considered negligible in 7 k . Plaice in 7 j is typically targeted by the Irish otter trawl fleet, which operate on sandy grounds off the southwest of Ireland, close to shore and this species is a small, but valuable component of the landings in a mixed fishery. Whereas, plaice in 7 h is mostly targeted by the beam trawl fleet, and some otter trawl, which operate close to the boundaries of other plaice stocks (ple.27.7.fg \& ple.27.7.e)(Figure 27.1).

To date no stock identification studies have been conducted on plaice in $7 \mathrm{~h}-\mathrm{k}$, which is on the southwestern margins of the species distribution, which is reflected in the reported landings that show high landings in adjoining stock areas, 27.7.e and 27.7.fg (Figure 27.1). There are no relevant tagging studies completed in this area. There is evidence in other areas to suggest that plaice is a highly mobile species, and therefore it is possible that ple.27.7.h-k is an extension of larger adjoining populations, but tagging and genetic would need to be completed to determine this (ICES, 2021).

### 27.2.2 Landings obligation

The EU multiannual plan (MAP) for stocks in the Western Waters and adjacent waters (EU, 2019) applies to bycatches of this stock. As of 2020, the EU landing obligation fully applied to plaice in divisions 7h-k. The landing obligation was phased in between 2016-2019 (Commission Delegated Regulations (EU) 2015/2438, 2016/2375, 2018/46, EU, 2015, 2016, 2018). A survivability exemption for plaice caught in ICES divisions $7 \mathrm{~h}-\mathrm{k}$ for vessels using beam trawls, with a maximum engine power of 221 kW , a maximum length of 24 metres, fishing within 12 nautical miles of the coast and with tow durations of no more than ninety minutes, and by vessels using beam trawls with an engine power of more than 221 kW , using a flip-up rope or benthic release panel (Commission Delegated Regulations (EU) 2020/2015). Additional survivability exemption available for plaice caught in pots, traps and creels in North-Western Waters (ICES subareas 5, 6 and 7) (Commission Delegated Regulations (EU) 2020/2015).

### 27.2.3 Management applicable to 2022 and 2021

TAC table 2022

| Species: | Plaice <br> Pleuronectes platessa |  |  | Zone: | $7 \mathrm{~h}, 7 \mathrm{j}$ and 7 k (PLE/7H)K.) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 7 | (1) | Preca | TAC |
| France |  | 14 | (1) |  | is Regulation |
| Ireland |  | 47 | (1) | Artic appl | gulation (EC) |
| Netherlands |  | 27 | (1) | Articl apply | gulation (EC) |
| Union |  | 95 | (1) |  |  |
| United Kingdom |  | 19 | (1) |  |  |
| TAC |  | 114 | (1) |  |  |

${ }^{(1)}$ Exclusively for by-catches. No directed fisheries for plaice are permitted under this TAC.

TAC table 2021

| Species: | Plaice <br> Pleuronectes platessa |  | Zone: | 7h, 7j and 7k (PLE/7HJK.) |
| :---: | :---: | :---: | :---: | :---: |
| Belgium | 4 | (1) | Preca | TAC |
| France | 8 | (1) | Artic | is Regulation |
| Ireland | 28 | (1) | Artic | gulation (EC) |
| The Netherlands | 16 | (1) | Artic | gulation (EC) |
| Union | 56 | (1) |  |  |
| United Kingdom | 11 | (1) |  |  |
| TAC | 67 |  |  |  |
| (1) | Exclusively for by-catches. No directed fisheries for plaice are permitted under this TAC. |  |  |  |

### 27.3 Data

### 27.3.1 Commercial catch data

Belgium, France, Ireland, Spain, and the United Kingdom (England) uploaded commercial catch data for 2021 to InterCatch (Figures 27.2 and 27.3). All submitted age samples are presented in Figure 27.4 and length samples in Figure 27.5. Although these samples are not used directly in the stock assessment they are used to determine a number of biological parameters which are used within the estimation of MSY proxy reference points.

Landings reported to InterCatch for this stock totalled 48 tonnes in 2021, which is a 9 tonne increase on 2020.

Discards submitted to InterCatch of plaice in divisions 7.h-k totalled 6 tonnes in 2021. As in 2020 there was again a decrease in commercial catch sampling due to COVID-19. The discard rates for this stock are highly variable over time (Figure 27.7), this variability may be driven by low
and variable sample sampling numbers. Due to this high interannual variability in estimated discards, an average discard rate of $42 \%$ was calculated from 2004-2019 and applied to the landings of the same period. To ensure that recent trends in discarding were captured, discard rates were calculated using the geometric means of the submitted discards for the five most recent years. These calculated to a mean discard rate of $50 \%$ in 2020 (2016-2020) and $49 \%$ in 2021 (20172021) (Figure 27.8).

### 27.3.2 Revisions

No revisions to previous years were submitted.

### 27.3.3 Survey indices

Seven fisheries-independent surveys were combined to model the first biomass index for plaice in this stock area. This modelled index was produced using VAST, which is a Vector Autoregressive Spatiotemporal model in R (Thorson et al., 2016). This model implements a spatial deltageneralized linear mixed model (delta-GLMM) to standardizing survey. VAST is spatially explicit model that predicts population density for all locations within a spatial domain, and then predicts derived quantities (i.e. biomass abundance) by aggregating population density across spatial domain while weighting density estimates by the area associated with each estimate.

The model was parametrised using haul level data from seven fisheries-independent surveys undertaken in the Celtic Sea (1997-2021) (Table 27.2). The coverage of these surveys varies in space and time, a full description of which can be found in Table 27.2 and Figure 27.9. The raw survey data were checked for quality (specifically, the estimated weights of the catch numbers-at-length were checked against the reported catch weights). For each valid haul, the catch weight, tow duration, tow position (midpoint), survey series and year were used as input values for the VAST model. The model was specified to have spatial autocorrelation but no temporal autocorrelation (i.e. years are independent). VAST can optionally estimate, and correct for, differences in catchability between the two survey series as there is a significant spatial overlap between the two surveys. The model first estimates the likelihood of occurrence and then the biomass using a gamma error distribution or the abundance using a lognormal error distribution. Historically none of these surveys were used to estimate abundances of plaice as individually they do not cover the full stock area, spatially/ temporally, and now of the surveys have been designed with this stock and species in mind. Vast offers a number of advantages over more traditional ways of estimating abundances. It has an ability to deal with gaps in survey coverage, and an ability to account for differences in catchability between surveys or vessels, providing an objective way to combine multiple indices even when the gear is not standardised.

The spatial domain was defined as 1000 knots, and implemented using k-means clustering to give knot positions proportional to sampling intensity (Thorson, 2019) (Figure 27.9). Residual diagnostics on the encounter probability appeared acceptable (Figure 27.10). Visualisation of the Pearson's residuals of positive catches (Figure 27.11a) and encounter probability (Figure 27.11b) show no strong patterns. These plots are the default output from the package, however in the future the presence/absence residuals should be revisited. The estimated survey biomass indices are presented in Table 27.3 and Figure 27.14, along with associated uncertainty. Visualisation of spatiotemporal variability in estimated log density of plaice in ICES divisions $7 \mathrm{~h}-\mathrm{k}$ (Figure 27.13), show distributional trends in areas of high abundance that mirror that of the known fishery, with high incidence of reported landings occurring in areas similar to the biomass from this VAST index, along the southwest coast of Ireland and the southwest coast of the UK. It is clear that these patches of high abundance spill over into adjoining stock area, plaice 7 fg , where landings are substantially higher than the plaice in $27.7 \mathrm{~h}-\mathrm{k}$.

### 27.3.4 Biological

A number of length-based parameters were required for the calculation of the new 'rfb' catch advice rule (ICES, 2020): mean length in observed catch $\left(\bar{L}_{y-1}\right)$, the length at first capture relative
 (length at $50 \%$ of mode) Lc.

The calculation of the ' rfb ' catch advice rule requires the calculation of $f$ which is the ratio of mean length $\left(\bar{L}_{y-1}\right)$ in the observed catch that is above the length of the first capture relative to the target length $(\mathrm{Lf}=\mathrm{m})$. The mean length in the observed catch was calculated by plotting the landings and discards data submitted to InterCatch over all years (2004-2021). The length of the first capture relative to the target length ( $\mathrm{LF}_{\mathrm{F}=\mathrm{m}}$ ) is calculated ( $\mathrm{L}_{\mathrm{F}=\mathrm{m}}=0.75^{*} \mathrm{~L}_{\mathrm{c}}+0.25^{*} \mathrm{~L}_{\infty}$ ). Length at first catch (length at $50 \%$ of mode)(Lc) was calculated from the landings and discards data submitted to InterCatch. $L_{c}$ was calculated for each year, but was found to be highly variable due to the variable and low sample number submitted for discards (Figures 27.15 and 27.16). Therefore, a mean of the time-series, 228.9 mm , was estimated as the $L_{c}$ of this stock (Figure 27.17) and used in the calculation of rfb. Similarly, the mean length in observed catch ( $\bar{L}_{y-1}$ ) was found to be was found to be highly variable due to the variable and low sample number submitted for discards (Figures 27.15 and 27.16). Therefore, a mean of the time-series, 297.3 mm , was estimated as the $\bar{L}_{\mathrm{y}-1}$ of this stock (Figure 27.18) and used in the calculation of rfb .
$L_{\infty}$ is calculated from the von Bertalanfy growth model. Samples available through DATRAS were used to calculate these length parameters. These samples were collected by three surveys, Irish ground fish survey (IGFS, 2004-2021), Irish anglerfish and megrim survey (IAMS, 20162021) and the French southern Atlantic bottom trawl survey (EVOHE, 2014-2021). Although none of these surveys are designed to capture the dynamics of this stock, they do provide the samples required to produce estimates of life history parameters. Only samples from 7j ( $\mathrm{n}=1648$ ) were used to calculate these parameters due to low sample size in $7 \mathrm{~h}(\mathrm{n}=11)$.

The FSA package in R (Ogle et al., 2022) was used to determine the starting values Ford-Walford $=($ vbStarts $\{F S A\})$ and to fit a von Bertalanffy growth curve was fit to the survey data for all areas combined, by bootstrapping a nonlinear regression (nls\{stats\}(R Core, 2020)). Due to the uneven sample size it was not possible to determine if these growth parameters vary between ICES divisions 7 j and 7 h . However, we could estimate the growth parameters for the whole stock as linf $=466.83 \mathrm{~mm}(\mathrm{SD} \pm 22.85), \mathrm{K}=0.18(\mathrm{SD} \pm 0.03), \mathrm{t} 0=-2.13(\mathrm{SD} \pm 0.31)$ (Figure 27.19). Residuals of model fitted considered acceptable (Figure 27.20).

### 27.4 Advice

### 27.4.1 Analyses of stock trends and potential status indicators

Advice is given based on trends in the VAST survey biomass index, The LBI-estimated values of the ratio $L m e a n L F=M$ are used to estimate exploitation status relative to the proxy MSY reference point, as described in the stock annex.
The advice for 2023 was set using the HCR ' rfb ' as outlined in the table below (ICES, 2022).

| Previous catch advice Ay 110 tonnes |  |
| :---: | :---: |
| Stock biomass trend |  |
| Index A (2020, 2021) | 587 |
| Index B (2017, 2018, 2019) | 457 |
| r: Index ratio (A/B) | 1.28 |
| Fishing pressure proxy |  |
| Mean catch length (Lmean=L2021) | 29.7 cm |
| MSY proxy length ( $\mathrm{LF}=\mathrm{m}$ ) | 28.9 cm |
| $f$ : multiplier for relative mean length in catches ( $\left.L_{\text {mean }} / L_{F}=\mathrm{M} 2020\right)$ | 1.03 |
| Biomass safeguard |  |
| Last index value (12021) | 150 tonnes |
| Index trigger value (Itrigger=Iloss×1.4) | 698 tonnes |
| b: multiplier for index relative to trigger (I2021/Itrigger; Itrigger = Iloss*1.4 = 1.05) | 1 |
| Precautionary multiplier to maintain biomass above Blim with 95\% probability |  |
| m : multiplier to maintain the biomass above $\mathrm{B}_{\text {lim }}$ to with $95 \%$ probability | 0.95 |
| Uncertainty cap (+20\%/-30\% compared to Ay, only applied if b $\geq 1$ ) Applied | 1.2 |
| Discard rate | 47\% |
| Catch advice for 2023 and 2024 (Ay $\times$ uncertainty cap) | 132 tonnes |
| \% advice change | + 20\% |

These results suggest that the relative fishing mortality is below the reference Fmsy proxy and the relative biomass is well-above the reference $\mathrm{BMSY}^{*} 0.5$ proxy. Therefore, the Precautionary Approach Buffer (PA Buffer) was not applied for the advice for this stock.

### 27.4.2 State of the stock

On the relative scale, the spawning biomass is estimated to have been since 2004 and has in remained high and stable in recent years (Figure 27.21). Estimated F was below Fmsy from 2004, and is now at the lowest point in the time-series.

### 27.4.3 Biological reference points

The table below summarises all known reference points for plaice in $27.7 \mathrm{~h}-\mathrm{k}$ and their technical basis. The LBI-estimated values of the ratio $L m e a n L F=M$ are used to estimate exploitation status relative to the proxy MSY reference point.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY <br> approach | MSY $B_{\text {trig- }}$ <br> ger proxy | 150 | Biomass index trigger value ( $l_{\text {trigger }}$ ), defined as $I_{\text {trigger }}=l_{\text {loss }} \times 1.4$, where $l_{\text {loss }}$ is the lowest observed historical biomass index value from 2005. In estimated metric tonnes from derived from VAST. | $\begin{aligned} & \text { ICES } \\ & (2022 a, b) \end{aligned}$ |
|  | $\mathrm{F}_{\text {MSY proxy }}$ | $\frac{L_{\text {mean }}}{L_{F=M}}=1$ | Relative value from $L B I$ analysis, assuming $M / K=1.5 . L_{F}=M$ is based on $L_{c}$ (length at $50 \%$ of modal abundance), which varies each year | ICES (2020) |
| Precautionary approach | $\mathrm{Blim}^{\text {m }}$ | Not defined |  |  |
|  | $\mathrm{B}_{\mathrm{pa}}$ | Not defined |  |  |
|  | $\mathrm{F}_{\text {lim }}$ | Not relevant |  |  |
|  | $\mathrm{F}_{\mathrm{pa}}$ | Not relevant |  |  |
| Management plan | $\mathrm{SSB}_{\text {MGT }}$ | Not relevant applicable |  |  |
|  | $\mathrm{F}_{\text {MGT }}$ | Not relevant applicable |  |  |

### 27.5 Recommendations for the next benchmark

This stock should be considered for the next SPiCT workshop to assess if it can be moved to a category 2 stock.

### 27.6 References

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Figure 27.1 The spatial distribution of plaice landings reported to the STECF fisheries-dependant information data call in 2016 (the last data year available), disaggregated by Member State (left) and gear (right). Note beam trawlers are described as beam and BT2, and otter trawlers are described as TR1 and TR2.


Figure 27.2. Plaice in Divisions 7.h-k. Landings and discards reported in InterCatch by country.


Figure 27.3. Plaice in Division 7.h-k. International landings reported in InterCatch by fleet and year.



Figure 27.4. Plaice in Division 7.h-k. Unraised landings (left) and discard (right) age distributions submitted to InterCatch.



Figure 27.5. Plaice in Division 7.h-k. Unraised landings (left) and discard (right) length distributions submitted to InterCatch.


Figure 27.6. Plaice in Division 7.h-k. Raw variable discard rates.


Figure 27.7. Plaice in Division 7.h-k. High inter-annual variability of discard rates supplied to InterCatch. Therefore, ICES discards are derived from estimated discard rates, 42\% (green line) from 2004-2019 (mean 2004-2019), 50\% (red line) in 2020 (geometric mean 2016-2020), 49\% (orange line) in 2021 (geometric mean 2017-2020).


Figure 27.8. Plaice in Division 7.h-k. Survey numbers per haul by year. Each point represents haul with a positive count shown as a circle and a zero as a ' + ' symbol. Circle diameter is proportional to the count. Colours denote the surveys.


Figure 27.9. Plaice in Division 7.h-k. The spatial area defined within the model in terms of latitude and longitude (top left), kilometres (top right) and knots (bottom).


Figure 27.10. Plaice in Division 7.h-k. Residual diagnostics showing predicted encounter probability against observed encounter probability (left) and the effective area occupied (right).


Figure 27.11(a). Plaice in Division 7.h-k. Spatiotemporal persons residuals (1) of encounter probability.


Figure 27.11(b). Plaice in Division 7.h-k. Spatiotemporal persons residuals (2) of encounter probability.


Figure 27.12. Plaice in Division 7.h-k. Spatiotemporal variability in estimated log density of plaice.


Figure 27.13. Plaice in Division 7.h-k. VAST estimated biomass in tonnes.


Figure 27.14. Plaice in Division 7.h-k.


Figure 27.15. Plaice in Division 7.h-k.


Figure 27.16. Plaice in Division 7.h-k. Lc.


Figure 27.17. Plaice in Division 7.h-k. Lmean.


Figure 27.18. Plaice in Division 7.h-k. Length (mm) versus age (dots) with superimposed best-fit von Bertalanffy growth function (black line) of all plaice in ICES divisions 27.7h and 27.7j available in Datras.


Figure 27.19. Plaice in Division 7.h-k. Residual plot (left) and histogram of residuals (right) of von Bertalanffy growth function (black line) on plaice in ICES divisions 27.7h and 27.7j available in Datras.

Table 27.1. Plaice in divisions 7.h-k. History of official landings by country and ICES estimated landings (tonnes).

| Year | BEL | FRA | IRL | UK | OTH | Official landings | ICES landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 0 | 60 | 321 | 286 | 0 | 667 |  |
| 1996 | 0 | 48 | 305 | 155 | 52 | 560 |  |
| 1997 | 0 | 69 | 344 | 138 | 0 | 551 |  |
| 1998 | 0 | 49 | 286 | 106 | 13 | 454 |  |
| 1999 | 45 | 0 | 299 | 82 | 1 | 437 |  |
| 2000 | 4 | 54 | 200 | 76 | 7 | 341 |  |
| 2001 | 27 | 50 | 160 | 73 | 3 | 313 |  |
| 2002 | 69 | 45 | 155 | 59 | 2 | 330 |  |
| 2003 | 20 | 32 | 127 | 56 | 6 | 241 |  |
| 2004 | 67 | 32 | 91 | 36 | 6 | 232 | 224 |
| 2005 | 32 | 20 | 90 | 28 | 0 | 170 | 167 |
| 2006 | 22 | 36 | 66 | 18 | 1 | 143 | 143 |
| 2007 | 7 | 30 | 72 | 20 | 13 | 141 | 124 |
| 2008 | 25 | 13 | 73 | 12 | 1 | 124 | 133 |
| 2009 | 1 | 44 | 72 | 32 | 0 | 149 | 143 |
| 2010 | <1 | 54 | 66 | 35 | 0 | 155 | 150 |
| 2011 | 4 | 58 | 72 | 44 | 0 | 178 | 172 |
| 2012 | 2 | 62 | 99 | 38 | 0 | 201 | 193 |
| 2013 | 0 | 49 | 52 | 40 | 0 | 141 | 137 |
| 2014 | 4 | 52 | 82 | 15 | 0 | 153 | 147 |
| 2015 | 0 | 60 | 25 | 18 | 0 | 103 | 107 |
| 2016 | 7 | 46 | 32 | 15 | 0 | 100 | 99 |
| 2017 | 11 | 53 | 42 | 10 | 0 | 116 | 114 |
| 2018 | 17 | 35 | 32 | 0 | 11 | 95 | 96 |
| 2019 | 6 | 17 | c | 9 | 2 | 34 c | 64 |
| 2020* | 7 | 12 | 12 | 7 | 1 | 39 | 39 |
| 2021* | 9 | 14 | 15 | 8 | <1 | 46 | 48 |

Table 27.2. Plaice in divisions 7.h-k. Surveys incorporated into VAST biomass index.

| Survey | Years | Quarters | Gear | Sources | Wing spread |
| :--- | :---: | :---: | :--- | :--- | :--- |
| IGFS | $2003-2021$ | 4 | Otter | DATRAS | Available at haul level |
| IAMS | $2003-2021$ | 1 |  <br> Beam | DATRAS | Available at haul level |
| EVOHE | $2003-2021$ | 4 | Otter | DATRAS | Available at haul level |
| WGCFS | $1997-2004$ | $1,2,4$ | Otter | CEFAS | Set to 21 m (average of other otter trawl surveys <br> in series) |


| SWBEAM | $2006-2021$ | 1 | Beam | DATRAS | Available at haul level |
| :--- | :--- | :--- | :--- | :--- | :--- |
| SWIBTS | $2003-2011$ | 4 | Otter | CEFAS | Set to 21 m (average of other otter trawl surveys <br> in series) |

Table 27.3. Plaice in divisions 7.h-k. Table abundance index - what is the SD log and SD mt.

| Year | Biomass estimate (tonnes) | SD_log | SD_mt |
| :---: | :---: | :---: | :---: |
| 1997 | 427.38 | 0.35 | 151.27 |
| 1998 | 189.42 | 0.49 | 92.74 |
| 1999 | 218.89 | 0.37 | 80.43 |
| 2000 | 127.41 | 0.53 | 67.64 |
| 2001 | 272.77 | 0.48 | 130.76 |
| 2002 | 145.11 | 0.54 | 78.71 |
| 2003 | 259.89 | 0.30 | 78.84 |
| 2004 | 243.33 | 0.33 | 81.28 |
| 2005 | 107.38 | 0.41 | 43.79 |
| 2006 | 123.99 | 0.36 | 44.58 |
| 2007 | 334.29 | 0.30 | 101.66 |
| 2008 | 368.39 | 0.27 | 100.58 |
| 2009 | 342.01 | 0.31 | 104.69 |
| 2010 | 339.91 | 0.28 | 93.98 |
| 2011 | 524.27 | 0.24 | 128.01 |
| 2012 | 501.62 | 0.25 | 124.23 |
| 2013 | 478.33 | 0.24 | 116.64 |
| 2014 | 760.40 | 0.23 | 171.45 |
| 2015 | 771.81 | 0.25 | 192.01 |
| 2016 | 885.08 | 0.20 | 180.68 |
| 2017 | 592.94 | 0.27 | 161.56 |
| 2018 | 459.78 | 0.25 | 114.38 |
| 2019 | 317.21 | 0.26 | 83.74 |
| 2020 | 474.79 | 0.28 | 134.20 |
| 2021 | 698.41 | 0.25 | 174.04 |

Table 27.4. Summary table for ple-7.jk assessment in input values.

| Year | Biomass index | Landings | Discards |
| :---: | :---: | :---: | :---: |
| 1995 |  | 667 |  |
| 1996 |  | 560 |  |
| 1995 |  | 667 |  |
| 1996 |  | 560 |  |
| 1997 | 427 | 551 |  |
| 1998 | 189 | 454 |  |
| 1999 | 219 | 427 |  |
| 2000 | 127 | 341 |  |
| 2001 | 273 | 313 |  |
| 2002 | 145 | 330 |  |
| 2003 | 260 | 241 |  |
| 2004 | 243 | 224 | 162 |
| 2005 | 107 | 167 | 121 |
| 2006 | 124 | 143 | 104 |
| 2007 | 334 | 124 | 90 |
| 2008 | 368 | 133 | 96 |
| 2009 | 342 | 143 | 104 |
| 2010 | 340 | 150 | 108 |
| 2011 | 524 | 172 | 124 |
| 2012 | 502 | 193 | 140 |
| 2013 | 478 | 137 | 100 |
| 2014 | 760 | 147 | 107 |
| 2015 | 772 | 107 | 77 |
| 2016 | 885 | 99 | 72 |
| 2017 | 593 | 114 | 83 |
| 2018 | 460 | 96 | 70 |
| 2019 | 317 | 64 | 47 |
| 2020 | 475 | 39 | 39 |
| 2021 | 698 | 48 | 46 |

## 28 Pollack in the Celtic Seas (ICES subareas 6 and 7)

## Type of assessment in 2022

The Celtic Sea and West of Scotland (subareas 6 and 7) Pollack stock is considered a Data-Limited Stock, classified by ICES WKLIFE II (ICES, 2012) as category 4.1.2. DCAC (Depletion-Corrected Average Catch) method is recommended to assess this stock, which is performed through the NOAA toolbox.

## ICES advice applicable to 2023

ICES advises that when the precautionary approach is applied, commercial catches should not exceed 3360 tonnes in 2023.

### 28.1 General

## Stock Identity

This section is not dedicated to a 'stock', it relates to a species in a wider region where data are available. The stock structure of Pollack populations in this ecoregion is not clear. ICES does not necessarily advocate that subareas 6 and 7 constitute a management unit for Pollack, and further work is required.

## Management applicable to 2022

The 2022 TAC for Pollack was set for ICES subareas 6 (and 5.a, b; international waters of 12 and 14) and 7 separately, Table 28.1.

The 2021 TAC for subarea 6 was 156 tonnes and uptake low at $41.3 \%$ and varied considerably between countries. France, which holds $48.0 \%$ of the TAC, only utilised $3.4 \%$ of their quota. The UK utilised $64.2 \%$ of the $36.5 \%$ TAC allocation, and Spain utilised none of their $1.3 \%$ TAC allocation. Ireland has $14.1 \%$ of the total TAC and had the largest uptake and exceeded their quota by $15.4 \%$ ( 30 tonnes as opposed to 26 tonnes).
In Subarea 7, which comprises the vast majority of landings and a TAC of 8012 tonnes, the uptake was also low at $21.3 \%$ and again varied considerably between countries. France, which holds the majority of the TAC allocation ( $67.0 \%$ ), only utilised $9.3 \%$ of this. The UK utilised $39.6 \%$ of its $22.7 \%$ TAC allocation, Ireland utilised $86.5 \%$ of its $7.1 \%$ TAC allocation, Belgium and Spain, which hold very low TAC allocations at $2.9 \%$ and $0.2 \%$, utilised $2.5 \%$ and $5.9 \%$ respectively.

## Fishery in 2021

## Landings

2087 tonnes of pollack were landed in 2021, $96.4 \%$ of which came from Subarea 7.
The nominal landings for ICES subareas 6 and 7 are shown in Tables 28.2 and 28.3 respectively.
For Subarea 6, there was an $58.3 \%$ increase in landings ( 76 tonnes) in 2021 compared to the landings in 2020 ( 48 tonnes), although this is still way below the TAC of 156 tonnes. The UK declared the highest landings (56.6\%) followed by Ireland (39.5\%). There was a 9\% decrease in landings
(2011 tonnes) for Subarea 7 in 2021 compared to 2020 ( 2210 tonnes). The UK had the highest landings ( $40.1 \%$ ) followed by France (29.5\%), Ireland (29.2\%), Belgium ( $0.3 \%$ ) and Spain ( $0.05 \%$ ).

## Landings by division

In 2021, $96.4 \%$ of catches came from Subarea 7 , with only $3.6 \%$ of landings derived from Subarea 6 and of those, over $99 \%$ came from division 6.a. In Subarea 7, the division with the highest proportion of landings derived from 7.e (42.7\%) followed by 7.g (17.1\%), 7.j (14.0\%), 7.h (13.6\%) and 7.f (7.9\%). Landings in divisions 7.a, b, c, d and k were negligible (4.7\%).

## Landings by gear

The majority of Pollack landings in the Celtic Sea ecoregions were caught by gillnets and trammel nets ( $51.9 \%$ ) followed by set lines ( $17.0 \%$ ), bottom trawlers ( $16.0 \%$ ), miscellaneous gears ( $10.9 \%$ ) and beam trawlers ( $2.8 \%$ ). When separated by subarea, the predominant gears landing pollack in Subarea 6 were bottom trawlers ( $57.3 \%$ ) followed by nets ( $28.6 \%$ ) and miscellaneous gears (10.2\%). In Subarea 7, nets had the highest landings ( $52.8 \%$ ) followed by lines (17.5\%), bottom trawlers ( $14.5 \%$ ), miscellaneous gears ( $10.9 \%$ ) and beam trawlers ( $2.9 \%$ ).

## Landings by quarter

Pollack are not historically targeted throughout the entire year, and are mainly targeted during the first quarter, which coincides with spawning. The breakdown of landings per quarter shows that the highest landings were in quarter $1(43.0 \%)$ followed by quarter 2 ( $23.6 \%$ ), quarter 3 ( $18.2 \%$ ) and quarter 4 ( $15.3 \%$ ) respectively.

## Discards

Discarding was negligible at less than a tonne in 2021.

## Landings uncertainty

Pollack is a known recreational fishing species, however; it is unknown as to the quantities exploited by recreational fisheries. A phone study conducted in France in 2011-2013 by Levrel et al. (2013) estimated that 3300 tonnes are landed annually through recreational fishing, 2274 tonnes of which are retained. Radford et al., 2018 further suggest that pollack landings may be similar to or above commercial landings. Work is currently being undertaken to provide recreational landing data.

### 28.2 Stock assessment

A DCAC (Depletion-Corrected Average Catch) method is used to estimate a yield likely to be sustainable (MacCall, 2009). Subarea 6 and 7 are run independently. For Subarea 6, six separate model runs using various parameters (Table 28.4) are conducted giving an average DCAC value plus an upper and lower $95 \%$ confidence interval and for Subarea 7 , nine separate model runs using various parameters (Table 28.5) are conducted giving an average DCAC value, plus an upper and lower $95 \%$ confidence interval.
The information provided for the assessment is insufficient to evaluate the exploitation and the trends of pollack in the Celtic Seas ecoregion. Commercial catches have declined since the late 1980s, and in 2021 are at historical lows for the combined Subarea 6 and 7.
The input data and parameters used for the assessment are detailed in Tables 28.4 and 28.5.

## 2022 Results

The average DCAC values (Figure 28.1) show that in both subareas 6 and 7, commercial landings are below the average DCAC by 66 tonnes in Subarea 6 and 1853 tonnes in Subarea 7. This suggests that yield in Subarea 6 could be increased up to 142 tonnes and 3864 tonnes in Subarea 7.

## Comparison with previous assessment

Table 28.6 compares the results with the previous year's assessment. The results are consistent with the range of DCAC values estimated when the method was previously applied.

## Uncertainties in assessment and forecast

The DCAC model relies solely on commercial catch data and does not include any biological or survey data that are available for this stock. The model also cannot estimate reference points. By construction, the DCAC method only uses long time-series of ICES landings and official landings. As the output is a smoothed value of the landings over the assessed time-series, the computations of DCAC are always similar to the previous year's results, even when recruitment or SSB fluctuate.

## Management considerations

TAC for Subarea 7 includes ICES Division 7.d, which is not in the remit of the Celtic Sea ecoregion. TAC set for both subareas 6 and 7 are not in line with the current estimates of catches and estimated sustainable yields, and therefore are not constraining.

## Management plan

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to demersal stocks including pollack in ICES subareas 6 and 7.

## Recommendations

This stock began the benchmark process in 2020 (ICES, 2021), but did not proceed past the data compilation phase. This was due to difficulties in quantifying the substantial recreational catch and concerns that at present, scientific surveys were unable to provide a representative abundance index. This was due to low Pollack catches in the surveys and the inshore, reef-based population that were not adequately sampled by existing surveys. Progress was made with regards to improved age and length sampling data from the main countries involved in the fishery so there is potential to push forward with data-limited methods utilising these data.

There is potential to collaborate with ICES Working Group on Recreational Fisheries Surveys (WGRFS) to improve on the existing UK data, and to bolster ongoing work in France and Ireland to provide data on the extent of their recreational fishing of Pollack. Although the recent benchmark did not lead to an improved assessment for Pollack, progress was made in identifying the data shortcomings that exist and potential routes forward. The need for an improvement on the existing DCAC assessment remains, and there is a commitment to progressing the data and assessment options. Based on continuing work this stock is due for benchmark in late 2022.

### 28.3 References

EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. Official Journal of the European Union, L 83: 1-17. http://data.europa.eu/eli/reg/2019/472/oj.

ICES. 2012. Report of The Workshop to Finalize the ICES Data-limited Stock (DLS) Methodologies Documentation in an Operational Form for the 2013 Advice Season and to make Recommendations on Target Categories for Data-limited Stocks (WKLIFE II), 20-22 November 2012, Copenhagen, Denmark. ICES CM 2012/ACOM:79. 46 pp.
ICES. 2021. Benchmark Workshop on selected stocks in the Western Waters in 2021 (WKWEST). ICES Scientific Reports. 3:31. 504 pp. https://doi.org/10.17895/ices.pub. 8137

Levrel H., Bellanger M., Le Goff R. and Drogou M. 2013. La pêche récréative en mer en France métropolitaine (Atlantique, Manche, Mer du Nord, Méditerranée). Résultats de l'enquête 2011-2013. http://archimer.ifremer.fr/doc/00162/27300/.

MacCall, A. D. 2009. Depletion-corrected average catch: a simple formula for estimating sustainable yields in data-poor situations. - ICES Journal of Marine Science, 66: 2267-2271.

Radford Z, Hyder K, Zarauz L, Mugerza E, Ferter K, Prellezo R, et al. 2018. The impact of marine recreational fishing on key fish stocks in European waters. PLoS ONE 13(9): e0201666. https://doi.org/10.1371/journal.pone. 0201666.

Table 28.1. 2022 TAC for Pollack in ICES subareas 6 (and 5.a, b; international waters of 12 and 14) and 7.

| Species: | Pollack <br> Pollachius pollachius | Zone:6; United Kingdom and international <br> waters of 5b; international waters of 12 <br> and 14 <br> (POL/56-14) |
| :--- | :--- | :--- |
| Spain | 2 | Precautionary TAC |
| France | 75 |  |
| Ireland | 22 |  |
| Union | 59 |  |
| United Kingdom | 156 |  |
| TAC |  |  |


| Species: | Pollack <br> Pollachius pollachius |  |  | Zone | $\begin{aligned} & 7 \\ & (\mathrm{POL} / 07 .) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 233 | (1) | Preca | TAC |
| Spain |  | 14 | (1) |  |  |
| France |  | 5372 | (1) |  |  |
| Ireland |  | 572 | (1) |  |  |
| Union |  | 6191 | (1) |  |  |
| United Kingdom |  | 1821 | (1) |  |  |
| TAC |  | 8012 |  |  |  |

${ }^{(1)}$ Special condition: of which up to $2 \%$ may be fished in United Kingdom, Union and international waters of 8a, 8b, 8 d and $8 \mathrm{e}(\mathrm{POL} / * 8 \mathrm{ABDE})$.

Table 28.2. Landings of Pollack in Subarea 6 as officially reported to ICES.

| Year | Belgium | Denmark | France | Germany | Ire- <br> land | Netherlands | Norway | Portugal | Spain | Sweden | UK | Total <br> Subarea 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 1 | - | - | - | - | - | - | - | - | - | 295 | 296 |
| 1951 | - | - | - | - | - | - | - | - | - | - | 484 | 484 |
| 1952 | - | - | - | - | - | 1 | - | - | - | - | 503 | 504 |
| 1953 | - | - | - | - | - | - | - | - | - | - | 422 | 422 |
| 1954 | - | - | - | - | - | - | - | - | - | - | 452 | 452 |
| 1955 | - | - | - | - | - | - | - | - | - | - | 566 | 566 |
| 1956 | - | - | - | - | - | - | - | - | - | - | 528 | 528 |
| 1957 | - | - | - | - | - | - | - | - | - | - | 547 | 547 |
| 1958 | . | - | - | 23 | - | - | - | - | - | - | 710 | 733 |
| 1959 | 1 | - | - | 6 | - | - | - | - | - | - | 607 | 614 |
| 1960 | 15 | - | - | - | - | - | - | - | - | - | 441 | 456 |
| 1961 | 1 | - | - | 1 | 125 | - | - | - | - | - | 259 | 386 |
| 1962 | 2 | - | - | 8 | 197 | - | - | - | - | - | 235 | 442 |
| 1963 | 6 | - | - | 2 | 204 | - | - | - | - | - | 320 | 532 |
| 1964 | 1 | - | - | 1 | 130 | - | - | - | - | - | 368 | 500 |
| 1965 | 1 | - | - | 1 | 402 | - | - | - | - | - | 496 | 900 |
| 1966 | 2 | - | - | - | 200 | - | - | - | - | - | 428 | 630 |
| 1967 | 1 | - | - | 1 | 263 | - | - | - | - | 1106 | 413 | 1784 |
| 1968 | 5 | - | - | 2 | 214 | - | 148 | - | - | 1012 | 500 | 1881 |
| 1969 | 1 | - | - | 4 | 282 | - | - | - | - | 1224 | 667 | 2178 |
| 1970 | 2 | - | - | 1 | 398 | - | - | - | - | 756 | 447 | 1604 |
| 1971 | 1 | - | - | 5 | 75 | - | - | - | - | 750 | 256 | 1087 |
| 1972 | 1 | - | - | 1 | 127 | - | - | - | - | 779 | 317 | 1225 |
| 1973 | 2 | - | - | - | - | - | - | - | - | - | 503 | 505 |
| 1974 | 6 | - | - | - | - | 3 | - | - | - | - | 359 | 368 |
| 1975 | < 0.5 | - | - | 1 | - | 1 | 4 | - | - | - | 393 | 399 |
| 1976 | 7 | - | - | - | - | 1 | - | - | - | - | 519 | 527 |


| Year | Belgium | Den- <br> mark | France | Germany | Ireland | Netherlands | Norway | Portugal | Spain | Sweden | UK | Total <br> Sub- <br> area 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | - | - | 196 | - | - | 1 | 2 | - | - | - | 493 | 692 |
| 1978 | - | - | 196 | - | - | - | 4 | - | - | - | 553 | 753 |
| 1979 | - | - | 310 | - | - | - | - | - | - | - | 350 | 660 |
| 1980 | - | - | 36 | - | - | - | - | - | - | - | 233 | 269 |
| 1981 | - | - | 342 | - | - | - | - | - | 55 | - | 185 | 582 |
| 1982 | - | $<0.5$ | 272 | - | - | - | - | - | 95 | - | 103 | 470 |
| 1983 | - | - | 331 | - | - | - | - | - | 86 | - | 148 | 565 |
| 1984 | - | - | 212 | - | - | - | - | - | 222 | - | 194 | 628 |
| 1985 | < 0.5 | - | 224 | 1 | - | - | - | - | 283 | - | 328 | 836 |
| 1986 | - | - | 145 | - | 223 | - | - | - | 2217 | - | 187 | 2772 |
| 1987 | - | $<0.5$ | 108 | - | 103 | - | - | - | 860 | - | 259 | 1330 |
| 1988 | - | $<0.5$ | 128 | - | 163 | - | - | - | 1925 | - | 221 | 2437 |
| 1989 | - | $<0.5$ | 111 | 1 | 103 | - | - | - | - | - | 179 | 394 |
| 1990 | - | - | 76 | - | 150 | - | 1 | - | - | - | 192 | 419 |
| 1991 | - | - | 31 | - | 145 | - | - | - | 4 | - | 189 | 369 |
| 1992 | - | $<0.5$ | 21 | - | 23 | - | - | - | < 0.5 | - | 203 | 247 |
| 1993 | - | - | 39 | - | 12 | - | - | - | - | - | 273 | 324 |
| 1994 | - | - | 34 | $<0.5$ | 26 | - | $<0.5$ | - | - | - | 276 | 336 |
| 1995 | - | - | 64 | 3 | 83 | - | - | - | - | - | 354 | 504 |
| 1996 | - | $<0.5$ | 29 | < 0.5 | 97 | - | 1 | - | - | - | 210 | 337 |
| 1997 | - | - | 14 | 1 | 69 | - | 2 | - | - | - | 162 | 248 |
| 1998 | - | - | 21 | - | 60 | - | - | < 0.5 | $<0.5$ | - | 147 | 228 |
| 1999 | - | - | - | - | 73 | - | 3 | - | $<0.5$ | - | 136 | 212 |
| 2000 | - | - | 11 | 2 | 62 | - | - | - | - | - | 116 | 191 |
| 2001 | - | - | 8 | - | 108 | - | - | - | - | - | 101 | 217 |
| 2002 | - | - | 9 | - | 26 | - | - | - | - | - | 96 | 131 |
| 2003 | < 0.5 | - | 3 | - | 88 | - | 1 | - | - | - | 111 | 203 |


| Year | Belgium | Denmark | France | Germany | Ire- <br> land | Netherlands | Norway | Portugal | Spain | Sweden | UK | Total <br> Subarea 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | < 0.5 | - | 2 | - | 68 | - | 1 | - | - | - | 65 | 136 |
| 2005 | - | - | 23 | - | 28 | - | - | - | - | - | 16 | 67 |
| 2006 | - | - | 3 | $<0.5$ | 31 | - | $<0.5$ | - | 4 | - | 5 | 42 |
| 2007 | - | - | 10 | $<0.5$ | 26 | - | 7 | - | - | - | 21 | 64 |
| 2008 | - | - | 8 | - | 21 | - | 1 | - | - | - | 23 | 54 |
| 2009 | - | - | 7 | - | 6 | - | $<0.5$ | - | - | - | 25 | 38 |
| 2010 | - | - | 6 | - | 34 | - | $<0.5$ | - | - | - | 39 | 80 |
| 2011 | - | - | 2 | - | 12 | - | 1 | - | - | - | 36 | 51 |
| 2012 | - | - | 2 | - | 10 | - | $<0.5$ | - | 2 | - | 33 | 48 |
| 2013 | - | - | 1 | - | 34 | - | $<0.5$ | - | - | - | 22 | 58 |
| 2014 | - | - | 1 | - | 25 | - | $<0.5$ | - | - | - | 18 | 44 |
| 2015 | - | - | $<0.5$ | - | 23 | - | $<0.5$ | - | - | - | 25 | 49 |
| 2016 | - | - | $<0.5$ | - | 44 | - | $<0.5$ | - | - | - | 31 | 76 |
| 2017 | - | - | $<0.5$ | - | 33 | - | $<0.5$ | - | - | - | 14 | 47 |
| 2018 | - | - | $<0.5$ | - | 22† | - | $<0.5$ | - | - | - | 29 | 51+ |
| 2019 | - | - | 3 | - | 22† | - | $<0.5$ | - | - | - | 27 | 51+ |
| 2020 * | - | - | $<0.5$ | - | 13 | - | - | - | - | - | 32 | 45 |
| 2021* | - | - | 2 | - | 29 | - | - | - | $<0.5$ | - | 43 | 75 |

*Preliminary commercial landings.
† Incomplete due to part of the data being unavailable under national GDPR clauses.

Table 28.3. Landings of Pollack in Subarea 7 as officially reported to ICES.

| Year | Belgium | Denmark | France | Germany | Ireland | Netherlands | Norway | Spain | UK | Total <br> Subarea <br> 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 93 | - | - | - | - | - | - | - | 375 | 468 |
| 1951 | 74 | - | - | 2 | - | - | - | - | 380 | 456 |
| 1952 | 80 | - | - | 10 | - | - | - | - | 336 | 426 |
| 1953 | 34 | - | - | - | - | - | - | - | 252 | 286 |
| 1954 | 17 | - | - | 4 | - | - | - | - | 365 | 386 |
| 1955 | 38 | - | - | - | - | - | - | - | 247 | 285 |
| 1956 | 67 | - | - | 1 | - | - | - | - | 155 | 223 |
| 1957 | 219 | - | - | 6 | - | - | - | - | 367 | 592 |
| 1958 | 342 | - | - | 17 | - | - | - | - | 233 | 592 |
| 1959 | 158 | - | - | 32 | - | - | - | - | 251 | 441 |
| 1960 | 317 | - | - | - | - | - | - | - | 267 | 584 |
| 1961 | 268 | - | - | - | 360 | - | - | - | 210 | 838 |
| 1962 | 367 | - | - | 1 | 369 | - | - | - | 170 | 907 |
| 1963 | 95 | - | - | - | 411 | - | - | - | 176 | 682 |
| 1964 | 299 | - | - | - | 342 | - | - | - | 194 | 835 |
| 1965 | 362 | - | - | - | 335 | - | - | - | 231 | 928 |
| 1966 | 456 | - | - | - | 438 | - | - | - | 175 | 1069 |
| 1967 | 417 | - | - | - | 474 | - | - | - | 202 | 1093 |
| 1968 | 214 | - | - | - | 508 | - | - | - | 167 | 889 |
| 1969 | 142 | - | - | - | 794 | - | - | - | 161 | 1097 |
| 1970 | 165 | - | - | 1 | 724 | - | - | - | 120 | 1010 |
| 1971 | 114 | - | - | - | 673 | - | - | - | 116 | 903 |
| 1972 | 142 | - | - | - | 1073 | - | - | - | 123 | 1338 |
| 1973 | 89 | - | - | - | - | 3 | - | - | 127 | 219 |
| 1974 | 299 | - | - | - | - | 13 | - | - | 223 | 535 |
| 1975 | 295 | - | - | - | - | 17 | - | - | 290 | 602 |
| 1976 | 339 | - | - | - | - | 4 | - | - | 421 | 764 |


| Year | Belgium | Denmark | France | Germany | Ireland | Netherlands | Norway | Spain | UK | Total <br> Subarea <br> 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 157 | 1 | 3569 | - | - | 1 | - | - | 465 | 4193 |
| 1978 | 186 | 21 | 5496 | 14 | - | 8 | - | - | 515 | 6240 |
| 1979 | 151 | 18 | 5119 | 76 | - | 1 | - | - | 696 | 6061 |
| 1980 | 237 | 7 | 5242 | - | - | 1 | - | 1 | 769 | 6257 |
| 1981 | 244 | - | 5814 | - | - | 3 | - | 23 | 780 | 6864 |
| 1982 | 154 | - | 4253 | - | - | - | - | 32 | 1022 | 5461 |
| 1983 | 167 | - | 6214 | - | - | - | - | 26 | 1045 | 7452 |
| 1984 | 207 | - | 3927 | - | - | - | - | 486 | 1100 | 5720 |
| 1985 | 269 | - | 3741 | - | - | - | - | 20 | 1022 | 5052 |
| 1986 | 241 | - | 4574 | - | 1335 | - | - | 17 | 1795 | 7962 |
| 1987 | 149 | - | 5213 | - | 848 | - | - | 19 | 2010 | 8239 |
| 1988 | 191 | - | 5211 | - | 1066 | - | - | 22 | 1740 | 8230 |
| 1989 | 145 | - | 3893 | - | 994 | - | - | 18 | 1487 | 6537 |
| 1990 | 133 | - | 4831 | - | 1066 | - | - | 26 | 1914 | 7970 |
| 1991 | 76 | - | 3211 | - | 1045 | - | - | 22 | 1962 | 6316 |
| 1992 | 62 | - | 2849 | - | 1014 | - | - | 19 | 1889 | 5833 |
| 1993 | 55 | - | 2325 | - | 1137 | - | - | 7 | 2135 | 5659 |
| 1994 | 94 | - | 2621 | - | 921 | - | - | 8 | 2391 | 6035 |
| 1995 | 88 | 2 | 2315 | - | 1107 | - | - | 4 | 2168 | 5684 |
| 1996 | 94 | - | 2684 | - | 1190 | 6 | - | 5 | 2519 | 6498 |
| 1997 | 99 | - | 2443 | - | 984 | 4 | < 0.5 | 7 | 2540 | 6077 |
| 1998 | 92 | - | 2375 | - | 886 | 1 | - | 11 | 2347 | 5712 |
| 1999 | 86 | - | - | - | 976 | - | 3 | 19 | 1703 | 2787 |
| 2000 | 71 | - | 2422 | - | 1069 | - | - | 5 | 1810 | 5377 |
| 2001 | 100 | - | 2515 | - | 1274 | - | - | 9 | 1987 | 5885 |
| 2002 | 117 | - | 2481 | - | 1308 | - | - | 17 | 1999 | 5922 |
| 2003 | 113 | - | 2284 | - | 1151 | - | - | 12 | 1788 | 5348 |


| Year | Bel- <br> gium | Denmark | France | Germany | Ireland | Netherlands | Norway | Spain | UK | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Subarea |  |  |  |  |  |  |  |  |  |  |
| 2 |  |  |  |  |  |  |  |  |  |  |

*Preliminary commercial landings.
$\dagger$ Incomplete due to part of the data being unavailable under national GDPR clauses.

Table 28.4. Input parameters for the six DCAC runs carried out for Pollack in Subarea 6.

|  | Run 1 | Run 2 | Run 3 | Run 4 | Run 5 | Run 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sum of catch | 6960 | 6960 | 6960 | 6960 | 6960 | 6960 |
| CV sumC | 0 | 0 | 0 | 0 | 0 | 0 |
| no of years | 36 | 36 | 36 | 36 | 36 | 36 |
| iterations | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| stdev M | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 |
| $\mathrm{F}_{\text {MSY }} / \mathrm{M}$ | 0.6 | 0.8 | 1 | 0.6 | 0.8 | 1 |
| stdev $\mathrm{F}_{\text {MSY }}$ to M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| distr $\mathrm{F}_{\text {MSy }}$ to M | lognormal | lognormal | lognormal | lognormal | lognormal | lognormal |
| $\mathrm{B}_{\mathrm{MSY}} / \mathrm{BO}$ | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 |
| stdev $\mathrm{B}_{\mathrm{MSY}} / \mathrm{BO}$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| up $\lim \mathrm{B}_{\text {MSY }} / \mathrm{BO}$ | 1 | 1 | 1 | 1 | 1 | 1 |
| low lim $\mathrm{B}_{\text {MsY }} / \mathrm{BO}$ | 0 | 0 | 0 | 0 | 0 | 0 |
| depletion delta $\Delta$ | 0.8 | 0.8 | 0.8 | 0.9 | 0.9 | 0.9 |
| stdev $\Delta$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| distr $\Delta$ | normal | normal | normal | normal | normal | normal |

Table 28.5. Input parameters for the 9 DCAC runs carried out for Pollack in Subarea 7.

|  | Run 1 | Run 2 | Run 3 | Run 4 | Run 5 | Run 6 | Run 7 | Run 8 | Run 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sum of catch | 174824 | 174824 | 174824 | 174824 | 174824 | 174824 | 174824 | 174824 | 174824 |
| CV sumC | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| no of years | 36 | 36 | 36 | 36 | 36 | 36 | 36 | 36 | 36 |
| iterations | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 | 10000 |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| stdev M | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 | 0.4 |
| $\mathrm{F}_{\text {MSY }} / \mathrm{M}$ | 0.6 | 0.8 | 1 | 0.6 | 0.8 | 1 | 0.6 | 0.8 | 1 |
| stdev $\mathrm{F}_{\text {MSY }}$ to M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| distr $\mathrm{F}_{\text {MSY }}$ to M | lognormal | lognormal | lognormal | lognormal | lognormal | lognormal | lognormal | lognormal | lognormal |
| $\mathrm{B}_{\text {MSY }} / \mathrm{BO}$ | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 |
| stdev $\mathrm{B}_{\text {Ms\% }} / \mathrm{BO}$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| up $\lim \mathrm{B}_{\text {Ms\% }} / \mathrm{BO}$ | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| low lim $\mathrm{B}_{\text {ms\% }} / \mathrm{BO}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| depletion delta $\Delta$ | 0.5 | 0.5 | 0.5 | 0.6 | 0.6 | 0.6 | 0.7 | 0.7 | 0.7 |
| stdev $\Delta$ | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| distr $\Delta$ | normal | normal | normal | normal | normal | normal | normal | normal | normal |

Table 28.6. Comparison of the 2020 DCAC assessment and previous DCAC results.

|  | Subarea 6 |  | Subarea 7 |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Landing (t) | Average DCAC | Landing (t) | Average DCAC |
| 2021 | 76 | 142 | 2011 | 3864 |
| 2020 | 48 | 143 | 2210 | 3906 |
| 2019 | 54 | 146 | 2165 | 3966 |
| 2018 | 63 | 148 | 2895 | 4010 |
| 2017 | 44 | 150 | 3260 | 4042 |
| 2016 | 74 | 152 | 4131 | 4063 |
| 2015 | 48 | 155 | 3740 | 4062 |
| 2014 | 44 | 156 | 5359 | 4020 |
| 2013 | 57 | 158 | 4468 | 3953 |



Figure 28.1. Pollack in subareas 6 and 7. The results of the depletion-corrected average catch (DCAC) assessment method as applied to commercial catch data since 1986. The grey box indicates the proxy for the maximum sustainable catch $\pm 95 \%$ confidence intervals.

# 29 Seabass (Dicentrarchus labrax) in divisions 4.b-c, 7.a, and 7.d-h (central and southern North Sea, Irish Sea, English Channel, Bristol Channel, and Celtic Sea) 

## Type of assessment

This is an update of the assessment accepted as the agreed methods to use at the benchmark workshop for the seabass: WKBASS (ICES, 2017-2018). The assessment is performed using the Stock Synthesis model implementation (SS3; Methot, 2000; 2011). The stock is treated as Category 1 with a full analytical assessment and forecast.

## ICES advice applicable to 2021

The ICES advice for management of seabass fisheries in 2021 is available in the ICES Advice released in 2019, and states that "when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, total removals in 2021 that correspond to the F ranges in the plan are between 1680 tonnes and 2000 tonnes".

## ICES advice applicable to 2022

ICES advises that when the MSY approach is applied, total removals in 2022 should be no more than 2216 tonnes. ICES notes the existence of a precautionary management plan, developed, and adopted by some of the relevant management authorities for this stock.

### 29.1 General

### 29.1.1 Stock definition and ecosystem aspects

Studies including tagging programmes and microchemistry are underway to provide information on the movement of sea bass and levels of mixing between stocks. Currently Atlantic stock identities are assumed to be as follows (ICES, 2012a,b):


In 2021, WGCSE included a session shared with WGBIE where a summary of some studies into stock identity summarized work from data storage tags, conventional tags, genetics, otolith microchemistry and larval dispersion models. The Working Groups recognized the complexity of these issues and considered that a stock identity workshop might be convened to allow relevant experts to consider these and any other relevant studies, and advise whether the existing stock boundaries remain appropriate.

### 29.1.2 Management

Historical management is described in the Stock Annex.

### 29.1.2.1 Management applicable from 2020 to 2022

In 2020, 2021 and 2022 the seabass fishery of stock bss.27.4bc7ad-h was prohibited, with derogations as shown in the simplified tables below for professional and recreational fishermen (season length, catch limits given per vessel for commercial and per fisher for recreational). See official regulations 2020, 2021, 2022 respectively for full details ${ }^{1,2,3}$

|  | Year | Demersal trawl ${ }^{[1]}$ | Seines ${ }^{[2]}$ | Hook and Lines ${ }^{[3]}$ | Fixed Gillnets ${ }^{[4]}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2020 | January 2020 and 04/01/2020 to 12/31/2020 | $\begin{aligned} & <520 \text { kg/2 months; } \\ & <5 \% \text { total catch/day } \end{aligned}$ | $\begin{aligned} & <520 \text { kg/2 months; } \\ & <5 \% \text { total catch/day } \end{aligned}$ | 5.7 t/year | 1.4 t/year |
|  | February-March 2020 | no landings | no landings | no landings | no landings |
| 2021 | January 2021 and 04/01/2021 to 07/31/2021 | $\begin{aligned} & <520 \mathrm{~kg} / 2 \text { months; } \\ & <5 \% \text { total catch/day } \end{aligned}$ | $\begin{aligned} & <520 \text { kg/2 months; } \\ & <5 \% \text { total catch/day } \end{aligned}$ |  |  |
|  | February-March 2021 | no landings | no landings | no landings | no landings |
|  | 08/01/2021 to 12/31/2021 | <380 kg/month | <380 kg/month |  |  |

[^15]
*January/April; May/June; July/August; September/October; November/December.

| 2020 | January-February-December 2020 | Recreational |
| :---: | :---: | :---: |
| 2021 | $01 / 03 / 2020$ to 30/11/2020 | 0 fish/day |
|  | January-February-December 2021 | 2 fish/day |
| 2022 | $01 / 03 / 2021$ to 30/11/2021 | 0 fish/day |
|  | January-February-December 2022 | 2 fish/day |

### 29.1.3 Fishery description

### 29.1.3.1 Total landings (official)

The history of the fishery is described in the Stock Annex. Table 1 and Figure 1 present official and total ICES landings. A large decrease in total landings was observed in 2014 due to poor weather conditions during winter and then from 2015 onwards due to management measures. Historically the bulk of the landings were made by the French fishery, but since implementation of management measures, landings are shared between French, UK and NL, and to a lesser extent Belgium. In 2021, 1275 tonnes were landed (official source): 613 t by UK, 385 t by France, 231 t by Netherland and 45 t by Belgium. Landings from France and the UK by gear are given in Figure 2. The COVID-19 pandemic didn't affect total commercial catches.

### 29.2 Data

### 29.2.1 Commercial landings

Landings are used for six fleets where selectivity is modelled (Table 2): fleet 1- UK bottom trawls and nets; fleet 2- UK lines; fleet 3- UK midwater pair trawls; fleet 4-French combined fleets; fleet 5- other countries plus UK gears not included in fleet 1, with selectivity based on fleet 4; and fleet 6 - recreational fisheries, where 2012 is the reference year. The source of information for the commercial fleets is the ICES database InterCatch. The time-series of recreational fisheries removals is calculated iteratively, so that fishing mortality remains constant and equal to the fishing mortality in 2012 over the period 1985-2014. After the implementation of the management in 2015, a multiplier is applied to recreational fishing mortality based on the severity of the measures (see chapter below). The landings are census data (EU logbooks and/or sales slips) from several sources:

1. Official statistics recorded in the ICES official landings database since the mid-1970s, with data from 1985 are used in this assessment.
2. French landings for 2000-2021 from a separate analysis of logbook, auction data and VMS data (SACROIS database) by Ifremer - extracted from the ICES database InterCatch.
3. Landings for Belgian vessels - extracted from the ICES database InterCatch.
4. Landings for Netherlands - extracted from the ICES database InterCatch Exception where a mistake was found in InterCatch, i.e. 2018 landings were updated for the 2020 assessment using official data.
5. UK landings by gear type recorded in official UK landings databases (historically and "InterCatch" database).

Details of the methodology used to calculate French and UK historical landings can be found in the Stock Annex.

### 29.2.2 Commercial length and age compositions

IBPBass2 (ICES, 2016) developed the Stock Synthesis model to include both the length and age compositions for the landings of fleets for which selectivity is estimated (Fleet 1: UK combined bottom trawl and nets -1985 onwards; Fleet 2: UK lines -1985 onwards; Fleet 3: UK midwater trawlers - 1985 onwards; Fleet 4: French combined gears -2000 onwards). Fitting to length composition data helps the estimation of length-based selectivity, whilst the age compositions (from application of age-length keys to length frequencies according to stratified sampling schemes) provide direct fitting of model estimates of catch-at-age. Since the length data are effectively being used twice, the length and age datasets are down weighted to avoid over-fitting of the data. The composition data for the fleets are given in the SS3 data file. Input sample sizes for the multinomial composition data are derived from numbers of fishing trips sampled, as proxy for effective sample size. The relative sample sizes between years are maintained in any reweighting.

### 29.2.2.1 Sampling rates

UK (England and Wales) sampling effort for length and age compositions by gear group are given in Table 3. The UK midwater trawl fleet landings were not sampled in 1997, 2013-2017 and since 2019 due to the small number of trips targeting seabass. This has negligible impact on the assessment as this UK métier represented only $1 \%$ of total seabass landings in 2013 and landed 2 t or less each year since 2014. In addition, Stock Synthesis will impute age distributions for missing years from the selectivity curve and landings.

Sampling of seabass in France also varied between gears (Table 4). Numbers of fish sampled decreased from 2015 due to the implementation of management measures and the fact that relatively few fish are now landed. The level of sampling was very low in 2019, with some of the main métiers, including lines and nets, not sampled. Level of sampling has nevertheless increased in 2020 and 2021.

The number of trips sampled in the UK is used as input in the stock assessment, with exception of a set number being attributed to UK midwater trawl and French fleet age composition. These numbers are then iteratively adjusted using the Francis method of weighting, reducing the disproportionate effect of the different datasets used.

### 29.2.2.2 Length composition estimates

Figure 3 and Table 5 give fleet-raised length compositions for all French gears combined. French numbers-at-length are available from 2000 onwards. The French fleet is the combination of several types of subfleets using a variety of fishing gears: pelagic trawlers, bottom trawlers, netters, liners, Danish seiners and purse-seiners (see details in Table 4). Figure 4, Figure 5 and Figure 6 give fleet-raised length compositions per UK métier used in the assessment (UK bottom trawls and nets; Lines; Midwater trawls).

### 29.2.2.3 Age composition estimates

The French age composition time-series from 2000 is from the application of an annual agelength key to the annual length composition of landings (Table 6).

Fleet-raised age compositions were obtained for UK fleets from 1985 onwards by application of age-length keys developed for the areas 4.bc, 7.d, 7.e\&h, and 7.a,f,g. The annual age compositions for the combined bottom trawl and nets fleet and the line fleet are given in Table 7 and Table 8, and the age compositions for the UK midwater pair trawl fleet since 1996 are given in Table 9.

### 29.2.3 Commercial discards

### 29.2.3.1 Discards and post-release mortality

Discarding of seabass below the MCRS occurs in most commercial fisheries to a variable extent. Previously, ICES advice sheets indicated overall international discard rates of only $5 \%$ by weight for the bss.27.4bc7ad-h stock. The WGCSE and WKBASS (ICES, 2017) showed that discard rates have typically been the highest in bottom otter trawls (OTB) and have increased following the introduction of additional management measures in 2015. Discards are now included in the assessment of this stock and in the absence of any data on discard survival, this has been assumed to be zero for all commercial fisheries. This has the potential to overestimate commercial fishing mortality, but the effect was initially expected to be small due to the low discard rates prior to 2015. This has changed in recent years, since the management measures have been implemented and discard rates are expected to increase in the short term as fishers adjust to take account of the changes, such as the increase in minimum conservation reference size from 36 cm to 42 cm .

Survival of fish discarded by commercial line vessels may be similar to survival of recreational angling releases (see next section), but work is needed to establish the typical gear, handling, and condition of fish to be released. Survival of seabass caught by trawls, seines, fixed or driftnets and longlines will depend on many factors including tow duration, soaking times, gear design, deep-hooking, and time on deck. There is need for studies on discard survival of seabass in different commercial fisheries.

### 29.2.3.2 Commercial discards data

Data sources for discards estimates and sampling design are described in the Stock Annex, with a summary of data from the UK and French on-board sampling programmes from 1985 to present given in Table 10, Table 11 and Table 12. Note that in the assessment, from 2015 onwards, discards from French observer data were replaced by logbook estimates, more realistic.

The observer estimates of annual discards by UK and French vessels from 2009 to 2014 was less than $5 \%$ of total landings. Between 2016 and 2021, the level of discarding observed in the French fleet increased compared to the previous period and varied between $10 \%$ and $22 \%$ (Table 12). This was mainly attributed to bottom trawlers where seabass is often a bycatch. French logbook data led to higher discards rate estimates of $23 \%, 38 \%, 59 \%, 53 \% 43 \%$ and $49 \%$ from 2016 to 2021. For the UK fleet a level of $9 \%$ was observed in $2016,6 \%$ in 2018 and up to $18 \%$ in 2019 , all mainly due to bottom trawlers (Table 11). In 2021, UK bottom trawlers discarded an estimated $19 \%$ of their catches. The level of sampling of the UK fleets since2017 has been low and so no raised estimate of discard length distribution was included in the assessment for these years.

French log book data from 2016 to 2021 showed that discard rates estimated from on-board sampling were much lower than those reported in logbooks. The logbook data provided estimates of discards of $155.6,270.9,456.4,374.6,313$ and 404 tonnes respectively, as opposed to the onboard sampling programme estimates of 152.7, 161.7, 34.2, 79.2, 3 and 13 tonnes for 2016, 2017, 2018, 2019, 2020, 2021 respectively. The increase in discards in 2018 may be explained by more restrictive management measures, but also by the fact that French fishers have been encouraged to report their discards in logbooks because of the landings obligation.

### 29.2.4 Recreational catches

The approach used for recreational catches is described in detail in the Stock Annex, but is briefly summarised here including the latest relevant data.

### 29.2.4.1 Recreational catches point estimates

Only a single year of recreational catches was available: 1440 t in 2012. This value of 1440 t was obtained by summing international recreational activities survey estimates for France, the Netherlands and the UK. It represented total removals through adding the retained fish and releases assuming a $5 \%$ post-release mortality. A composite length-frequency distribution was generated for recreational removals from the same survey data, with a post-release mortality of $5 \%$ applied to the release component.

### 29.2.4.2 Recreational removals time-series reconstruction

F for the recreational fishery was assumed to be constant prior to the introduction of management measures in 2015. Limited survey data were available after the implementation of management measures at the time of the benchmark in 2017-2018, so no reliable catch estimates existed. As a result, a method was developed for estimating the impact of combinations of the MCRS, season length and bag limits on removals by recreational fishing. A multiplier was derived from 2012 catches in terms of numbers of fish for the recreational F that related to the reduction in catch due to management. This corresponded to multipliers of 0.821 in 2015, 0.282 in 2016 and 2017, 0.191 in 2018, 0.312 in 2019 and 0.464 for 2020, 2021 and 2022 (Table 13).

Since completion of the benchmark in 2018, further surveys have been conducted in the UK, France, Belgium, and the Netherlands that provide estimates of recreational catches of seabass. However, these surveys have been done in different years, using different methods, and have different associated biases. It is not obvious how best to combine the data for use in the assessment and would represent a significant departure from the current approach. Hence, this should be done as part of the next benchmark and peer-reviewed to ensure it robustness. As a result, the current approach will continue to be used until the next benchmark and recreational catches included on the issue list.

### 29.2.5 Biological data

All parameters for growth, weight, maturity, natural mortality and ageing error were as described in the Stock Annex.

### 29.2.6 Survey data used in assessment

### 29.2.6.1 Pre-recruit surveys in UK

An inshore trawl survey in autumn in a major bass nursery area in the Solent (7.d English coast) provides abundance indices-at-ages 2 to 4 for the stock assessment (Figure 7). Data are available from 1982, although there are intermittent years when the survey did not take place. The index calculation was updated in 2020 after a rigorous quality assessment was conducted (Table 14 and Table 15). The Stock Annex provides details of this survey (SBTS, G9863) and of some other prerecruit survey series not considered appropriate by previous WGs and IBPBass for inclusion in the assessment.

### 29.2.6.2 Pre-recruit surveys in France

Similar surveys have been done by Ifremer along the coast of France since 2014 to provide insight into French seabass nurseries areas and pre-recruit dynamics. The new time-series is not
available to WGCSE, but is expected to provide additional information on seabass age groups 03 to be considered for inclusion in the assessment. In the Channel, the survey takes place in the Seine estuary and preliminary indices are available from 2017. The survey is expected to continue until 2023 under a European Maritime and Fisheries Fund (EMFF) program (NOURDEM) and the working group will encourage its continuation after the index generated is reviewed and if it provides valuable information and supports the assessments. Indices will be tested during the next benchmark.

### 29.2.6.3 Channel Groundfish survey FR-CGFS

The French Channel Groundfish survey (FR-CGFS, G3425) has been carried out in October each year since 1988. It provides swept-area indices of seabass abundance in the Eastern Channel (7.d) together with length compositions. The swept-area indices are given in Table 16. Details of the survey can be found in Coppin et al. (2002) and sampling stations shown in Figure 8. The majority of seabass are caught in the coastal waters of England and France. The original time-series finished in 2014 as a new vessel was used for the survey from 2015. The new time-series now includes seven years of data, so may be considered for inclusion in the assessment at the next benchmark.

### 29.2.7 Commercial landings per unit of effort

Following the recommendation from WKBASS (ICES, 2018) the French LPUE index is now calculated by modelling the zeros and non-zeros values using a delta-GLM approach (see Stock Annex for details). Confidence interval calculated through a bootstrap estimation are presented in Table 17 and Figure 9, with the updated LPUE series used in the assessment.

### 29.2.8 Other relevant data

None.

### 29.3 Stock assessment

### 29.3.1 Model structure and input data / parameters for update assessment

The assessment was conducted using Stock Synthesis (Methot, 2000; 2011), using version 3.24u (Methot, 2011). The structure and input data / parameters of the SS3 model are summarized below and details are available in previous sections.

### 29.3.1.1 Model structure

- Temporal unit: annual based data (landings, discards, survey and commercial tuning indices, age and length frequencies).
- Spatial structure: One area.
- Sex: Both sexes combined.


### 29.3.1.2 Fleet definition

Six fleets defined: 1. UK bottom trawls, nets; 2. UK lines; 3. UK midwater trawls; 4. French fleets (combined); 5. Other (other countries and other UK fleets combined); 6. Recreational fisheries.

### 29.3.1.3 Landings and discards

Annual landings in tonnes from 1985 to final assessment year for the five fleets from ICES subdivisions $4 . b$ and c, 7.a, d-h. Recreational catch for 2012 with the time-series from 1985 to present iteratively reconstructed conditioned on the 2012 estimated value of 1440 t .
Discards in tonnes for fleet 1 (UK bottom trawls, nets) from 2002 and fleet 4 (French) from 2009.

### 29.3.1.4 Abundance indices and compositional data

Channel Groundfish Survey in 7.d in autumn (France), 1988 to 2014 (FR-CGFS, G9527): total swept-area abundance index and associated length composition data (Table 16). Input CV for survey is 0.60 for 1988-1990 and 0.30 for 1991 to 2014. First three years of composition data are excluded due to sampling levels and high uncertainty in the data. For remaining years, number of stations with seabass is used as input effective sample size of compositional length data.

Cefas Solent Autumn bass survey (7.d) (SBTS, G9863): years 1986 to 2009, 2011, 2013 to present, for ages $2-4$. Selection was fitted as a function of length using a double normal model, with minimum and maximum ages specified as 2 and 4 in the age selection function (Table 15).

French LPUE: as updated every year.

### 29.3.1.5 Fishery landings age composition data

The age bin is set from 0 to 15 with a plus group for ages 16 and over. Age compositions for fleets are expressed as fleet-raised numbers-at-age, although they are treated as relative compositions in SS3. Year range for UK bottom trawls/nets and UK lines is 1985 to present; UK midwater pair trawl is 1996 to 2018 (no samples for 1997, 2013-2014, 2016-2017, 2019-present); French is all fleets from 2000 to present.

### 29.3.1.6 Fishery landings length composition data

The length bin is set from 4 to 100 cm by 2 cm intervals. Length compositions for fleets are expressed as fleet-raised number-at-length. Year range for UK bottom trawls/nets is 1985 to present; UK lines 1985 to present; UK midwater pair trawl 1985 to 2012 (no samples for 1997, 2013-2019-present); French all fleets from 2000 to present.

### 29.3.1.7 Model assumptions and parameters

Table 18 summarises key model assumptions and parameters. Other parameter values and input data characteristics are defined in the SS3 control file BassIVVII.ctl, the start.SS file, the forecast file Forecast.SS and the data file BassIVVII.dat.

### 29.3.1.8 Incorporation of recreational fishery catch estimates

2012 catch input and F multipliers on all other years to iteratively estimate the full time-series of recreational catches; calculations for the final assessment run are given in Table 21.

### 29.3.1.9 Final update assessment: diagnostics

The likelihood components ( $\log L^{*}$ Lambda) for the update SS3 assessment are given below:

| Likelihood components | Likelihood |
| :--- | :--- |
| TOTAL | 710.35 |
| Catch | $8.26 \mathrm{e}-013$ |
| Equilibrium catch | 0.0259 |
| Survey | -41.21 |


| Likelihood components | Likelihood |
| :--- | :--- |
| Discards | 26.24 |
| Length compositions | 398.4 |
| Age compositions | 302.79 |
| Recruitment | 24.06 |
| Forecast Recruitment | 0.032851 |
| Parameter soft bounds | 0.01808 |

A range of model outputs and diagnostics are given in Figures 13-33.
Good correspondence was found between the observed and fitted length and age compositions for each fleet (Figures 16-28). However, the fit to the French length compositions was poorer in 2014-2019 and the fit to French age composition was variable throughout the time-series. Some diagonal residual patterns are noted in the commercial age compositions indicating some problems in fitting extreme variations in recruitment.

Any smearing of age estimates from a strong year class into neighbouring weak ones could be responsible for year-class residuals in the UK age compositions that are apparent in the first half of the series. The age error vector included in the model helps to accommodate this in the fit to age compositions. The combined fit of the age and length composition data aggregated over the series look good (Figure 19 and Figure 25).

The survey abundance indices both fisheries-independent and fishery-dependent are fitted reasonably well (Figure 26, Figure 27 and Figure 28).The UK Solent autumn survey is characterised by a large variability with outliers present in the model fit (Figure 26). The model fits closely to the low indices for recent years because there are few fishery composition data for estimating these recent year classes.

The model is able to predict recruitment deviations back to around the 1974 year class due to the strong year classes captured in the data in the early years (Figure 29) allowing a longer term perception of recruitment dynamics. Recruitment is highly variable with no evidence of a reduction in average recruitment at the lower SSB values (Figure 29) although this perception is affected by the imposition of a steepness value of 0.999 for the fitted Beverton-Holt stock-recruit curve. Sensitivities to differing values for this parameter carried out during the benchmark workshops found that likelihoods progressively worsened as the steepness value was reduced.

### 29.3.2 Analytical retrospective analyses

Retrospective analysis with a five-year peel was carried out for the calculation of the Mohn's rho. This analysis shows that there is some evidence of a retrospective pattern, see table below and Figure 30, for recruitment, SSB and fishing mortality. However, the retrospective bias is within the tolerance threshold accepted by ICES ( -15 to +20 ) for SSB and fishing mortality, there has been no tolerance threshold set for recruitment.

| Spawn-stock biomass | 0.162 |
| :--- | :---: |
| Fishing mortality (ages 4-15) | -0.126 |
| Recruitment (age 0) | 0.236 |

The model is sensitive to the recent change in selectivity due to management measures where a block change in the selectivity and retention parameter estimates where introduced for data proceeding 2015.

### 29.3.3 Final update assessment: long-term trends

The time-series of estimates of numbers-at-age, combined recreational and commercial $\mathrm{F}_{(4-15)}$, are given in Table 19 and Table 20, and a summary of SSB, recruitment, F and commercial and recreational catch are given in Table 21 and Figure 31. These series are based on the final SS3 update run with 2021 set as the final year.

A sharp increase in F between 2011 and 2013 is generated because the assessment model interprets that landings were maintained despite a rapid decline in biomass. This may be a plausible scenario where aggregations or predictable migration routes of seabass can be targeted, and it is possible for fisheries to maintain landings as total stock size declines, and hence inflict an increasing fishing mortality rate. The F has since decreased in-line with sharp reduction in catches due to the discontinuation of the French midwater trawl and the implementation of additional management measures. In 2021 F has not changed compared to 2020, reflecting the latest management measures. SSB increased slightly which may have resulted from the management measures in place since 2015, and some above average recruitment events since 2013 as described below.

WGCSE has concluded that strong year classes in 1989 and some subsequent years caused a rapid increase in biomass throughout the stock area, and landings and fishing mortality in the commercial fishery also increased. The combined commercial and recreational fishery $F$ was well above Fmsy prior to 2015. Recruitment has been declining since the mid-2000s, and has been poor since 2008, however the recruitment estimated for 2013, 2014, 2016, 2018 and 2019 is above the long-term geometric mean of 15 652. Uncertainties in the assessment are explored in a subsequent section.

### 29.3.4 Comparison with previous assessments

With the addition of the 2021 data and the updated French LPUE, the time-series of recreational catch was updated to remain consistent with the assumption of a constant F for the period 1985 to 2014 and an F multiplier reduction for 2015 to present (Figure 32).

With these changes included in the update assessment, the perception of the stock has remained largely unchanged. The spawning-stock biomass, fishing mortality and recruitment estimated in 2021 when compared with the recent assessment (Figure 33) are very close, well within the 95\% confidence intervals.

### 29.3.5 The state of the stock

The marked increase in biomass in the 1990s was driven by the very strong 1989 year class and a number of subsequent strong year classes. The biomass prior to this was declining during a period of poor recruitment, and the recent decline in biomass also coincided with a period of
poor recruitment, but under conditions of higher F than estimated for the 1980s. The stock has been characterised by periods of poor recruitment in the 1980s and since 2008. These periods of poor recruitment have a major impact on biomass, which is exacerbated by any increase in F. Total biomass changes more quickly than SSB, due to the time taken for fish to reach maturity. An increasing trend in biomass was estimated since 2018, which may have resulted from the management measures in place to restrict catches since 2015 and the occurrence of a number of just above average recruitment events since 2013.

The period of increasing SSB in the 1990s and early 2000s also coincided with expansion of the stock in the North Sea. The enhanced productivity and geographic range of the stock at this time also coincided with a period of elevated sea temperatures (see WGCSE and Stock Annex for UK inshore sea temperature trends in relation to seabass recruitment).

The assumption of a constant recreational fishing mortality over time implies that recreational harvests were a much larger fraction of total fishery removals in the 1980s compared with the 2000s onwards (Figure 10). It is likely that in the 1970s or earlier, seabass was primarily the target of recreational fishing.

### 29.4 Biological reference points

The fishing pressure and biomass PA and MSY reference points defined by WKBASS (2018) were updated during WGCSE 2019 due to the inclusion of additional 2018 data and a new LPUE series which changed the perception of the stock. The details of the calculations of the new reference points are given in the Stock Annex with reference points given below.

In 2021, ICES revised the basis $\mathrm{F}_{\mathrm{pa}}$ for all stocks to use $\mathrm{F}_{\mathrm{p} .05 \text {, as a result the value is now } 0.203 \text {. All }}$ other reference points were unchanged.

| Reference points | Value |
| :--- | :--- |
| Precautionary Approach |  |
| Blim | 10313 |
| $\mathrm{~B}_{\text {pa }}$ | 14439 |
| Flim $^{\text {F }} \mathrm{pa}$ | 0.254 |
| MSY Approach | 0.203 |
| FMSY |  |
| FMSY lower | 0.1713 |
| FMSY upper | 0.142 |
| MSY Btrigger | 0.1713 |

### 29.5 Short-term predictions

Inputs for a short-term forecast are given in Table 22, and their derivation is explained below.

### 29.5.1 Recruiting year-class strength

Recruitment estimates for seabass were below average from 2008 to 2012 (Table 21). Since recruitment is at a low level since 2008 the working group agreed to only include 2010 to 2019 (10 years) for the geometric mean recruitment for the forecast (10 104thousand), this was also identified and advised by the ADG in 2019. This is summarised in the text table below:

| Year class | SS3 (age 0) | GM 2010-2019 |
| :---: | :---: | :---: |
| 2019 | 12 587thousand |  |
| 2020 | 10 104thousand |  |
| 2021 | 10 104thousand |  |
| 2022 | 10 104thousand |  |

### 29.5.2 Numbers of fish in 2022

These were derived from the update Stock Synthesis run with final year set at 2021. The numbers for ages 0-2 in 2022 were adjusted using the ratio of LTGM to SS3 values for 2020-2022 age 0 as explained above and in Stock Annex.

### 29.5.3 F-at-age vectors

Status quo F-at-age for the commercial fishery was taken as the average F-at-age as estimated from the last three years derived from the update Stock Synthesis run with final year set at 2021. This approach was taken to allow for the change in selectivity associated with the implementation of new management measures (Table 13).

The recreational F vector was estimated in a similar way using the average of the last three years, however the final $\mathrm{F}_{\text {bar }}$ was scaled using F multipliers on the 2012 F in Table 13 taking into account the management measures in place. For the intermediate year (2022), this was a nine-month open season with a two bag limit and a MCRS of 42 cm . Additional years' Fs were scaled to keep the F of the recreational fleet proportional to the F of the commercial fleet as in the intermediate year 2021.

### 29.5.4 Weights-at-age

Mean weights-at-age in the stock were taken from the Stock Synthesis output. The commercial fishery weights for 2021 were derived as a weighted mean of the values for French and UK fleets given in the Stock Synthesis output, using the model estimates of catch numbers for the two fleets as weighting factors. The annual weights-at-age for any fleet are time-invariant, as they are derived from length-at-age derived from von Bertalanffy growth curve parameters, with selectivity applied where appropriate. Length at Amax (30 years) was estimated as 80.26 cm .

### 29.5.5 Maturity ogive

The proportion mature at-age is the length-based ogive applied to the length-at-age distributions around the input VB growth curve, calculated within Stock Synthesis.

### 29.5.6 Detailed short-term forecast output at status quo $F$

A detailed short-term forecast is given in Table 22 assuming that $F$ in 2022 and 2023is the average of 2019-2021 from the assessment for the commercial fleet, and for the recreation fleet the partial F used is that described in Section 29.5.3.

Fishing in 2021 at the same fishing mortality as in 2018-2020 for the commercial fleet, and with the current two bag limit for nine months for the recreational fleet, an SSB of 12153 t is predicted in 2022, increasing from 11619 t in 2021 . With the same fishing effort in 2022 the SSB would go up to 12371 t . There is uncertainty in the forecast, as the actual rate of decline in population abundance in recent years is likely to be more uncertain than indicated by the SS3 model confidence limits. Also, the effect of the final package of technical and other management measures for seabass in 2015 to present are not fully known at this stage, and information will be needed on their implementation and effectiveness before their impact on fishing mortality can be ascertained. The assumption of constant recreational F is also untested.

### 29.5.7 Management options

WGCSE provides management options in which F multipliers are applied proportionally to commercial and recreational F-at-age (Table 23). In reality, fisheries managers may wish to allocate the combined forecasted landings in any way considered appropriate, and this would imply differing F-multipliers applied to each fishery.

The management options table includes options for a number of different scenarios and include F of 0.153 calculated by reducing FMSY by the stock size relative to MSY Btrigger for combined commercial and recreational fishing. This would provide combined commercial and recreational catches of 2554 tonnes. This would be an increase of $15.3 \%$ compared to the advice for 2022 . The allocation between commercial and recreational fisheries depends on the balance of controls applied on recreational and commercial fishing in 2021 and 2022.

With zero F in 2023, SSB is expected to increase from 12835 t in 2023 to 14753 tin 2024. Therefore, it is now possible to achieve $\mathrm{B}_{\mathrm{pa}}$ or MSY $\mathrm{B}_{\text {trigger }}$ (both 14439 tonnes) in this time period.

### 29.6 Uncertainties and bias in assessment and forecast

### 29.6.1 Landings and discards data

Historical landings of small-scale national fisheries not supplying EU logbooks or sales slips are known to be inaccurate. IBPBass ran the Stock Synthesis model with and without additional UK landings for nets and lines estimated from a separate Cefas logbook scheme, and found this had relatively little impact on stock trends or fishing mortality, but rescaled the biomass and recruitment due to the additional catch. However, if the extent of non-reporting is changing over time, for example to develop track record in the possible event of a future TAC, then bias will be introduced in the assessment trends.

Discard rates are low in most fisheries other than trawls. Estimates of discards are available only from the early 2000s, but do not cover all fisheries, are imprecise, and are only included for some fleets in the assessment. The overall discard rate by weight is thought to be less than $5 \%$ before the implementation of management measures, increasing in recent years. Nonetheless, a timeseries of discards at-length or -age is needed for all fleets if the impact of technical measures to improve selectivity is to be evaluated as part of any future bass management.

### 29.6.2 Fishery composition data

The ability to fit selectivity patterns for defined groups of fishery métiers, and to detect changes in selectivity, depends on collection of adequate numbers of independent, representative samples of length and age to sufficiently characterise the length or age compositions of the selected métier groups. What constitutes "sufficient" is impossible to define without simulation studies to examine relationship between precision of input data and the precision of estimates required for management.

The absence of length composition data for French fisheries prior to 2000 is a serious deficiency in the model preventing any evaluation of changes in selectivity that may have occurred, for example due to changes in the mix of gear types. The numbers of trips of each métier group sampled on shore in France and the UK has varied widely over time, and in the UK has declined substantially since the 2000s. In France, sampling effort has also been very low in recent years and now appears to only cover trawls when a large portion of the fleet is composed of other nets and lines. Currently, there are no composition data supplied by Netherlands and Belgium.
ICES has developed extensive advice on establishing statistically-sound sampling designs for estimating fishery length and age compositions and discard quantities (see reports of ICES Workshops on Practical Implementation of Statistically Sound Catch Sampling Programmes (WKPICS1-3, available on ICES website). Stratified random sampling of fishing vessels or harbours may lead to low sample sizes for species such as seabass for which large fractions of the total catches may be taken in relatively small numbers of fishing trips. The cost-benefit of expanding the sampling in vessel or harbour strata where most seabass landings are recorded,
without compromising statistical sampling design, should be investigated. The next benchmark should evaluate if sampling is currently sufficient to support continued application of Stock Synthesis fitting selection parameters to fishery composition data.

### 29.6.3 Recreational fishery harvests

Current assessments accommodate an estimate of recreational fishery landings in the assessment and forecasts based on landings from 2012 (ICES, 2016; 2018). This a crude approach based on surveys for only a year or two in France, UK, and the Netherlands, and leads to an assumption of constant recreational fishing mortality over time. Recreational catches have been observed to vary significantly over time in other fisheries, so this assumption of constant mortality is unlikely to be true.

Since completion of the benchmark in 2018, further surveys have been done of recreational catches, but it is not obvious how best to combine the data for use in the assessment and would represent a significant departure from the current approach. Hence, this should be done as part of the next benchmark and peer-reviewed to ensure its robustness. At this stage, it will also be necessary to review the assumptions made related to changes in recreational fishing mortality and selectivity over time, and consider splitting recreational catches by country and/or fate of fish.

Release rates are expected to increase due to bag limits and increases in MCRS that are in place or planned. Current studies of post-release mortality are limited, and more studies are needed to develop a better understanding of the fate of released fish given the high incidence of catch-and-release practices in sea angling for seabass.

### 29.6.4 Surveys

The Channel Groundfish Survey included in the assessment provides data on a wider range of sizes and ages than the Cefas Solent survey, though with a steeply domed size selection pattern. From 2015 onwards, Ifremer no longer used the scientific vessel "Gwen Drez" which was replaced by the larger vessel "Thalassa". A calibration exercise was carried out in 2014 to assess the effect of this change to a larger vessel. WGCSE noted a concern that coverage of the coastal waters of 7.d could be altered by the use of this new vessel (the size of the vessel may prevent fishing as close to the coast as was possible with the previous vessel). The results of the calibration exercise were evaluated and it was found that the series could not be extended beyond 2014 and that a new series would need to be created from 2015 onward. This new dataseries is still to be considered for inclusion in the assessment.

The Cefas pre-recruit surveys are now reduced to just the Solent autumn survey, with the Solent spring and the Thames survey having been removed by previous benchmark assessments as being unsuitable. Recruitment estimates for the most recent years are heavily dependent on the Solent survey, and it is important to maintain this series. However, there is a need for information on recruitment trends in other areas, as it cannot be assumed that the Solent index will in the long term represent overall recruitment patterns throughout areas 4 and 7 . There are several studies that have demonstrated spatial and temporal variation in abundance of seabass in estuaries in the UK, France, and Ireland. It would be useful to review and, if possible, include additional time-series at the next benchmark and consider a developing a broader survey of nursery habitats for all species. For information, such a survey is conducted in France from 2017 onwards in the Seine estuary through the project Nourdem (and also in the Bay of Biscay in the Loire estuary from 2016 and in the Gironde estuary from 2019). In 2022, a time-series of six years will be available, also a juvenile seabass index should be tested in the model. The trawl used was
developed for catching seabass. An average of 5500 bass are caught during the survey, allowing a calculation of robust index.

### 29.6.5 Commercial LPUE indices

The reliance of the assessment on the Solent and Channel trawl surveys is a potential source of bias because they cover only a part of the stock range, and the selectivity is heavily skewed towards young bass. This is of principle concern in establishing the current rate of decline in spawning-stock biomass and associated trends in fishing mortality. In the absence of relative abundance indices for older bass from surveys or commercial fishing vessels covering the range of the stock, it is difficult for the model to fit the recent stock trends and fishing mortality. Statistical modelling of French LPUE data by vessel and rectangle by Laurec and Drogou (WGCSE 2015, Annex 3, WD 07) is used in the assessment.

Analyses of UK commercial fishery LPUE, based on averaging across ICES rectangles where the bulk of seabass catches have been recorded, was presented to IBPNEW in 2012 (ICES, 2012a). There were divergent trends between fleets where seabass are typically a bycatch, and mainly under 10 m vessels where increased targeting has probably been occurring using lines and nets. Future development of UK LPUE indices together with equivalent French data would require careful evaluation of potential for LPUE of each fleet to track abundance. Further analyses on the validity of the French LPUE as an index of abundance should also be considered, especially in light of the current restrictions of fishing activities.

### 29.6.6 Stock structure and migrations

The assessment treats all seabass in 4.b,c and 7.a,d-h as a single biological stock, but there can be extensive migrations. For example, migrations are expected to occur between the south of the area and the Bay of Biscay (which is treated separately in the WGBIE group), or between the North Sea and the Channel, there is also strong site fidelity (Pawson et al., 2008) resulting in a high proportion of tagged fish being recaptured at the same coastal location, even in subsequent years after migrations to offshore spawning sites. Immature seabass may remain close inshore, and exploitation of young fish in coastal waters ( $<6$ nautical miles offshore) may be predominantly by inshore fleets of that country. Mature fish originating from coastal waters of the UK, France or Netherlands or other countries may become increasingly vulnerable to offshore pelagic pair trawlers fishing mainly on mature fish during December to April. These spatial, ontogenetic patterns may lead to complex responses of length and age compositions to previous fishery catches of each country and fleet. This could potentially be addressed using spatial structuring in Stock Synthesis, but the data demands would increase substantially. Both the UK (England) and France have studies underway to improve knowledge of seabass movement and mixing.

### 29.6.7 Biological parameters

The maturity ogive used in the assessment was derived from sampling from the 1980s onwards. There has been no coordinated sampling across the full range of the stock in recent years to determine if the current ogive is still valid. Sporadic recent sampling has suggested that seabass may be spawning at sizes smaller than recorded historically (see Stock Annex). This would alter the Fmsy and could also be associated with changes in growth parameters. Mean length-at-age in UK samples remained more or less constant over several decades of sampling, but this analysis needs updating. Changes in growth, or inappropriate growth parameters, will lead to bias in fitting length-selectivity parameters to the French fishery and survey data.

# 29.6.8 Intermediate year fishing mortality and catch levels for forecasts 

Measures introduced by the UK government and EU commission to reduce fishing mortality toward Fmsy have the potential to affect the short-term forecast assumptions for this stock. Table 22 and Table 23 provide a detailed short-term status quo forecast and a range of management options from the forecast run.

### 29.7 Recommendations

### 29.7.1 Management considerations

Seabass in this stock are characterised by slow growth, late maturity and low natural mortality of adults, which imply the need for comparatively low rates of fishing mortality to avoid depletion of spawning potential in each year class. Productivity of the stock is affected by extended periods of enhanced or reduced recruitment, which appear to be related to changes in sea temperature. Warm conditions facilitate northward penetration of seabass in the North Sea and Northeast Atlantic and enhance the growth and survival of young fish in estuarine and other coastal nursery habitats. A period of above-average sea temperatures and enhanced recruitment between 1989 and the mid-2000s generated a large increase in biomass and a geographic expansion. Increased abundance and a lack of a TAC or other means to control fishing outside of nursery areas stimulated a growth of fisheries and markets for seabass. Many small-scale artisanal fisheries, especially line fishing and some forms of netting, have developed a high seasonal dependency on seabass, and there is also a significant recreational fishing mortality in inshore waters. The behaviour of seabass, forming predictable aggregations for spawning and moving close inshore to feed at other times of year, increase their vulnerability to exploitation by offshore and inshore fisheries. Increased targeting of seabass has resulted in a progressive increase in fishing mortality above values considered appropriate to achieve Fmsy. The combination of increasing fishing mortality and environmental conditions causing poor recruitment since 2008 appears responsible for a continuous decline in biomass since 2010. Catches appear to be declining in fisheries where seabass is mainly a bycatch, but some other fisheries such as netting in the UK appear to be expanding and may be exploiting known seasonal migration routes and local aggregations of fish despite a more widespread contraction of the population.

Careful management of fishing pressure on seabass is needed to prevent SSB declining to such an extent that the stock's ability to produce strong recruitment in more favourable environmental conditions is impaired. Since 2013, the European Commission developed a package of management measures to promote recovery of the stock. This resulted in emergency measures to stop the offshore pelagic trawl fishery on spawning aggregations between January and April 2015, bag limits for recreational fishing, and an increase the MCRS to 42 cm . Further measures to restrict catches without resorting to a TAC have been implemented. Any management measures applied to commercial and recreational fisheries should take into account the need for collection of data to demonstrate the effectiveness of the measures, and the ability to enforce the measures adequately.
ICES advice in 2004 recommended that "implementation of 'input' controls, preferably through technical measures aimed at protecting juvenile fish, in conjunction with entry limitations into the offshore fishery in particular should be promoted", and that "any consideration of catch limitation (output control) would need to take into account that seabass are a bycatch in mixed fisheries to a various extent, depending on gear and country; this incites discarding and should be avoided". This form of advice has re-occurred in subsequent ICES advice for seabass.

WGCSE notes that protection of juvenile fish through technical measures is good to improve the fishery selectivity and increase the number of seabass that are able to spawn at least once, but this is probably not enough to ensure a sufficient decrease in F. Protection of juveniles already exists to an extent already through designation of 37 UK seabass nursery areas where certain types of fishing on seabass is prevented annually or seasonally. However, catching and discarding of seabass by trawlers fishing close to nursery areas remains an issue. Data available to WGCSE indicate that discarding is mainly by otter trawlers using $80-90 \mathrm{~mm}$ mesh in or near areas where juvenile bass are most abundant, for example in UK coastal waters of the eastern Channel. Improvements to fishery selectivity to successfully achieve a large reduction in fishing mortality on pre-spawning fish without increasing discarding would require changes to gear designs, which could have a strong spatial management component.

Entry limitation can prevent an increase in effort but will not decrease F to the extent needed, unless existing licences are withdrawn. The occurrence of seabass as a small bycatch in many fisheries raises the problem of this becoming a "choke species" if vessel catch limits are introduced and seabass fall under the landings obligation.

ICES also previously advised that "Management of seabass fisheries needs to take into account the distinctive characteristics and economic value of the different fisheries. Seabass is of high social and economic value to the large inshore artisanal fleets and to sea angling and other recreational fishing that contribute substantially to local economies". Data from France indicated that the first sale value of the high-volume and lower quality catches of seabass caught by pelagic trawlers targeting offshore spawning fish during December to March had been up to three times lower per kg than for smaller volume sales of higher-quality fish from métiers fishing inshore (Drogou et al., 2011). However, there is at present insufficient information to accurately evaluate the total economic value and impact of seabass fisheries beyond just the first sale value and covering direct incomes from sales and direct as well as indirect and induced costs, employment and added value generated downstream. The interrelationship between markets for wild caught and farmed seabass also needs to be evaluated. A number of studies on the economic value of recreational sea fisheries have been conducted in recent and these demonstrate high levels of spend into national economies years (e.g. Armstrong et al., 2013; Roberts et al., 2017; Hyder et al., 2018a). Marine recreational fisheries in Europe has been shown to have a total economic impact of 10.5 billion and support almost 100,000 jobs in Europe (Hyder et al., 2017; 2018a), but this cannot be easily split between individual species.

No bio-economic scenarios are available at present to appreciate the effect of management measures for seabass, based on economic considerations, and work is urgently needed in this area. The importance of seabass to recreational fisheries, artisanal and other inshore commercial fisheries and large-scale offshore fisheries in different regions means that resource sharing is an important management consideration that has implications for the type of scientific evidence needed. A number of studies have shown that recreational catches of bss.27.4bc7ad-h represent around one quarter of the total catch (Armstrong et al., 2013; Hyder et al., 2018b; Radford et al., 2018).

The effects of targeting of offshore spawning aggregations of seabass in the English Channel and Celtic Sea are poorly understood, particularly how the fishing effort is distributed in relation to mixing of fish from different nursery grounds or summer feeding grounds in the UK, France and other countries, given the strong site fidelity of seabass.

The current stock structure assumptions are pragmatic and need further evaluation. The seabass population in coastal waters of the Republic of Ireland is currently considered as a separate stock, although it extends into at least one of the ICES divisions defining the $4 . \mathrm{bc}$ and $7 . \mathrm{a}, \mathrm{d}-\mathrm{h}$ stock. Further studies are needed to determine if the seabass in Irish coastal waters are indeed functionally separate, or if they also mix with the other stock during spawning time and contribute
to commercial catches on the offshore spawning grounds. Moreover, the Bay of Biscay is also currently considered as a separate stock although tagging programme indicates some exchange with the area 4 and 7 stock studied assessed by WGCSE.

As bass is, at present, a non-TAC species, there is potential for continued displacement of fishing effort from other species with limiting quotas. The fisheries on seabass have grown in the 1990s and 2000s due to good recruitment, and new markets have been established, competing with farmed bass. Fishing mortality gradually increased over time and was above Fmsy for many years. With the stock in decline measures were introduced to prevent the risk of stock collapse. Currently the likelihood of collapse remains high unless strong year classes are produced again and the management measures in place are continued and remain flexible to improving the fishery selection pattern, and limit total fishing mortality across all ages of seabass.

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Table 1. Bss.27.4bc7ad-h: Annual landings from 4b\&c and 7a,d-h (official landings per country and total ICES estimates).

| Year | Belgium | Denmark | Germany | France | UK | Netherlands | Channel Is. | Total | Total ICES |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0 | 0 | 0 | 620 | 105 | 0 | 18 | 743 | 994 |
| 1986 | 0 | 0 | 0 | 841 | 124 | 0 | 15 | 980 | 1319 |
| 1987 | 0 | 0 | 0 | 1226 | 123 | 0 | 14 | 1363 | 1980 |
| 1988 | 0 | 18 | 0 | 714 | 173 | 8 | 12 | 925 | 1239 |
| 1989 | 0 | 2 | 0 | 675 | 192 | 2 | 48 | 919 | 1161 |
| 1990 | 0 | 0 | 0 | 609 | 189 | 0 | 25 | 824 | 1063 |
| 1991 | 0 | 0 | 0 | 726 | 239 | 0 | 16 | 982 | 1227 |
| 1992 | 0 | 0 | 0 | 721 | 148 | 0 | 36 | 906 | 1186 |
| 1993 | 0 | 1 | 0 | 718 | 230 | 0 | 45 | 994 | 1255 |
| 1994 | 0 | 1 | 0 | 593 | 535 | 0 | 49 | 1178 | 1371 |
| 1995 | 0 | 1 | 0 | 801 | 708 | 0 | 69 | 1579 | 1835 |
| 1996 | 0 | 1 | 0 | 1703 | 563 | 8 | 56 | 2331 | 3022 |
| 1997 | 0 | 1 | 0 | 1429 | 561 | 1 | 74 | 2066 | 2620 |
| 1998 | 0 | 2 | 0 | 1363 | 488 | 48 | 79 | 1980 | 2390 |
| 1999 | 0 | 1 | 0 | na | 685 | 32 | 108 | 826 | 2670 |
| 2000 | 0 | 5 | 0 | 1522 | 407 | 60 | 130 | 2124 | 2407 |
| 2001 | 0 | 2 | 0 | 1619 | 458 | 77 | 80 | 2236 | 2500 |
| 2002 | 0 | 1 | 0 | 1580 | 627 | 96 | 73 | 2377 | 2622 |
| 2003 | 154 | 1 | 0 | 1903 | 586 | 163 | 84 | 2891 | 3459 |
| 2004 | 159 | 1 | 0 | 1883 | 617 | 191 | 159 | 3010 | 3731 |
| 2005 | 206 | 1 | 0 | 1937 | 512 | 327 | 220 | 3203 | 4430 |
| 2006 | 211 | 2 | 0 | 2116 | 736 | 308 | 23 | 3396 | 4377 |
| 2007 | 178 | 1 | 0 | 2075 | 873 | 376 | 18 | 3521 | 4064 |
| 2008 | 187 | 0 | 0 | 1506 | 934 | 380 | 20 | 3027 | 4107 |
| 2009 | 174 | 0 | 0 | 2904 | 801 | 395 | 15 | 4288 | 3889 |
| 2010 | 216 | 4 | 0 | 3441 | 879 | 399 | 14 | 4952 | 4562 |
| 2011 | 152 | 2 | 0 | 2688 | 928 | 395 | 17 | 4183 | 3858 |
| 2012 | 154 | 3 | 0 | 2492 | 946 | 376 | 12 | 3982 | 3987 |
| 2013 | 146 | 4 | 2 | 2868 | 841 | 370 | 12 | 4243 | 4137 |
| 2014 | 148 | 1 | 1 | 1322 | 1080 | 253 | 11 | 2816 | 2682 |
| 2015 | 40 | 0 | 0 | 1113 | 701 | 218 | 9 | 2081 | 2066 |
| 2016 | 23 | 0 | 1 | 545 | 551 | 156 | 24 | 1300 | 1295 |
| 2017 | 22 | 0 | 0 | 423 | 438 | 132 | 12 | 1027 | 984 |
| 2018 | 18 | 0 | 0 | 297 | 432 | 172 | 11 | 931 | 948 |
| 2019 | 19 | 0 | 0 | 309 | 411 | 209 | 22 | 970 | 972 |
| 2020* | 24 | 0 | 0 | 387 | 521 | 218 | 0 | 1150 | 1042 |
| 2021* | 45 | 0 | 0 | 385 | 613 | 231 | 1 | 1275 | 1126 |

Source: Official Landings Statistics. 2020 and 2021 provisional data; Total ICES, from InterCatch database.

Table 2. Bss.27.4bc7ad-h: Landings for the country / fleet components included separately in the assessment model.

| Year | Fleet 1: UK <br> Trawls, nets | Fleet 2: UK Lines | Fleet 3: UK pelagic trawlers | Fleet 4: France combined gears | Fleet 5: Other countries and gears | Fleet 6: <br> RecFish |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 70 | 30 | 1 | 870 | 23 | 1820 |
| 1986 | 84 | 33 | 2 | 1180 | 19 | 1639 |
| 1987 | 96 | 18 | 0 | 1840 | 25 | 1493 |
| 1988 | 129 | 30 | 8 | 1028 | 44 | 1384 |
| 1989 | 141 | 29 | 7 | 917 | 67 | 1279 |
| 1990 | 128 | 18 | 22 | 849 | 47 | 1150 |
| 1991 | 152 | 60 | 14 | 971 | 29 | 1056 |
| 1992 | 105 | 23 | 8 | 1001 | 49 | 1080 |
| 1993 | 146 | 62 | 1 | 979 | 68 | 1258 |
| 1994 | 354 | 154 | 0 | 786 | 76 | 1509 |
| 1995 | 424 | 169 | 4 | 1057 | 181 | 1689 |
| 1996 | 308 | 128 | 87 | 2395 | 104 | 1696 |
| 1997 | 335 | 119 | 71 | 1984 | 111 | 1605 |
| 1998 | 241 | 121 | 85 | 1773 | 170 | 1548 |
| 1999 | 274 | 148 | 220 | 1843 | 185 | 1548 |
| 2000 | 236 | 53 | 52 | 1805 | 261 | 1601 |
| 2001 | 263 | 58 | 97 | 1883 | 199 | 1685 |
| 2002 | 361 | 75 | 110 | 1825 | 251 | 1782 |
| 2003 | 353 | 65 | 127 | 2471 | 443 | 1863 |
| 2004 | 380 | 72 | 131 | 2604 | 544 | 1908 |
| 2005 | 353 | 59 | 68 | 3161 | 789 | 1905 |
| 2006 | 359 | 119 | 11 | 3259 | 629 | 1874 |
| 2007 | 413 | 166 | 37 | 2771 | 677 | 1871 |
| 2008 | 514 | 163 | 17 | 2750 | 663 | 1878 |
| 2009 | 486 | 147 | 9 | 2649 | 598 | 1852 |
| 2010 | 452 | 183 | 42 | 3236 | 649 | 1751 |
| 2011 | 462 | 143 | 98 | 2526 | 629 | 1605 |
| 2012 | 564 | 185 | 49 | 2610 | 579 | 1440 |
| 2013 | 530 | 191 | 39 | 2871 | 506 | 1222 |
| 2014 | 751 | 236 | 1 | 1303 | 391 | 1008 |
| 2015 | 440 | 199 | 0 | 1110 | 317 | 690 |
| 2016 | 305 | 210 | 2 | 547 | 231 | 209 |
| 2017 | 125 | 147 | 0 | 442 | 270 | 206 |
| 2018 | 160 | 267 | 0 | 313 | 208 | 150 |
| 2019 | 134 | 259 | 1 | 329 | 249 | 274 |
| 2020 | 190 | 306 | 0 | 409 | 137 | 453 |
| 2021 | 228 | 361 | 0 | 413 | 124 | 489 |

Table 3. Bss.27.4bc7ad-h: Sampling of commercial fishery landings of otter (A.), pelagic midwater trawls (A.), lines (B.) and nets (B.) for length and age in the UK (England and Wales). Nsamp is the number of landings (trips) sampled; Nfish is the number of fish measured.

| A. <br> Year | UK Otter trawl Age |  | Length <br> Nsamp | Nfish | Landings (t) | UK Pelagic/midwater |  |  | Nfish | Landings (t) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Age <br> Nsamp |  |  | Nfish | Length Nsamp |  |  |
| 1985 | 45 | 235 |  | 15 | 225 | 27 | 3 | 44 | 2 | 43 | 1 |
| 1986 | 18 | 216 | 28 | 2591 | 24 |  |  |  |  | 2 |
| 1987 | 41 | 421 | 54 | 1181 | 41 | 4 | 42 | 1 | 589 | 0.02 |
| 1988 | 23 | 257 | 23 | 1298 | 65 | 2 | 64 | 2 | 1684 | 8 |
| 1989 | 63 | 531 | 44 | 1595 | 80 | 4 | 126 | 4 | 1451 | 7 |
| 1990 | 63 | 883 | 48 | 773 | 67 | 8 | 19 |  |  | 22 |
| 1991 | 92 | 983 | 32 | 731 | 39 | 12 | 125 | 1 | 1490 | 14 |
| 1992 | 69 | 699 | 17 | 398 | 41 | 2 | 50 | 2 | 220 | 8 |
| 1993 | 118 | 1219 | 38 | 836 | 80 | 9 | 39 |  |  | 1 |
| 1994 | 182 | 1927 | 113 | 3925 | 125 |  |  | 1 | 127 | 0.3 |
| 1995 | 28 | 529 | 66 | 1995 | 162 |  |  | 1 | 19 | 4 |
| 1996 | 49 | 660 | 39 | 1041 | 122 | 1 | 41 | 3 | 392 | 87 |
| 1997 | 59 | 1660 | 52 | 2445 | 140 | 1 | 49 |  |  | 71 |
| 1998 | 28 | 676 | 39 | 1442 | 133 | 20 | 95 | 4 | 167 | 85 |
| 1999 | 24 | 379 | 46 | 1216 | 138 | 12 | 382 | 9 | 770 | 220 |
| 2000 | 92 | 759 | 42 | 1814 | 133 | 23 | 847 | 14 | 2463 | 52 |
| 2001 | 45 | 851 | 49 | 2152 | 141 | 3 | 58 | 5 | 691 | 97 |
| 2002 | 54 | 523 | 47 | 1454 | 161 |  |  | 4 | 545 | 110 |
| 2003 | 48 | 512 | 45 | 1418 | 207 | 15 | 459 | 4 | 744 | 127 |
| 2004 | 33 | 361 | 31 | 1295 | 173 | 8 | 161 | 5 | 522 | 131 |
| 2005 | 35 | 498 | 31 | 2432 | 181 | 3 | 149 | 2 | 299 | 68 |
| 2006 | 15 | 252 | 17 | 810 | 160 | 1 | 43 | 1 | 100 | 11 |
| 2007 | 44 | 385 | 21 | 903 | 173 | 1 | 20 | 3 | 355 | 37 |
| 2008 | 37 | 580 | 32 | 2151 | 196 | 6 | 409 | 8 | 1283 | 17 |
| 2009 | 24 | 1184 | 13 | 807 | 175 | 8 | 317 | 6 | 625 | 9 |
| 2010 | 25 | 360 | 28 | 1312 | 150 | 7 | 153 | 3 | 376 | 42 |
| 2011 | 25 | 577 | 49 | 1903 | 137 | 3 | 103 | 4 | 463 | 98 |
| 2012 | 18 | 182 | 41 | 751 | 157 |  |  | 1 | 199 | 49 |
| 2013 | 15 | 289 | 23 | 859 | 125 |  |  |  |  | 39 |
| 2014 | 14 | 164 | 22 | 523 | 104 |  |  |  |  | 1 |
| 2015 | 28 | 377 | 39 | 1277 | 100 | 1 | 4 | 1 | 4 | 1 |
| 2016 | 19 | 256 | 90 | 527 | 52 |  |  |  |  | 2 |
| 2017 | 38 | 510 | 128 | 915 | 51 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 43 | 263 | 43 | 492 | 28 | 1 | 15 | 1 | 33 | 0 |
| 2019 | 30 | 105 | 89 | 686 | 15 | 0 | 0 | 0 | 0 | 1 |
| 2020 | 47 | 90 | 47 | 251 | 27 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 51 | 151 | 51 | 376 | 36 | 0 | 0 | 0 | 0 | 0 |


| A. <br> Year | UK Otter trawl |  |  |  |  | UK Pelagic/midwater |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age <br> Nsamp | Nfish | Length <br> Nsamp | fish | Landings (t) | Age <br> Nsamp | Nfish | Length <br> Nsamp | Nfish | Landings (t) |
| B. | UK Lines |  |  |  |  | UK Nets |  |  |  |  |
|  | Age |  | Length |  | Landings (t) | Age |  | Length |  | Landings (t) |
| Year | Nsamp | Nfish | h Nsamp | Nfish |  | Nsamp | Nfish | Nsamp | Nfish |  |
| 1985 | 53 | 395 | 519 | 285 | 30 | 34 | 332 | 15 | 181 | 43 |
| 1986 | 60 | 496 | 631 | 894 | 33 | 18 | 251 | 18 | 1132 | 61 |
| 1987 | 92 | 313 | 369 | 557 | 18 | 37 | 528 | 44 | 1321 | 55 |
| 1988 | 66 | 538 | 853 | 1325 | 30 | 37 | 584 | 40 | 1397 | 64 |
| 1989 | 249 | 652 | 26 | 310 | 29 | 49 | 469 | 45 | 1248 | 60 |
| 1990 | 281 | 918 | 822 | 260 | 18 | 24 | 207 | 11 | 456 | 61 |
| 1991 | 346 | 1468 | 83 | 963 | 60 | 57 | 481 | 30 | 583 | 113 |
| 1992 | 418 | 2905 | 5111 | 2077 | 23 | 40 | 281 | 28 | 1248 | 64 |
| 1993 | 287 | 1787 | 723 | 1426 | 62 | 127 | 1141 | 94 | 1686 | 66 |
| 1994 | 212 | 1616 | 6155 | 3783 | 154 | 146 | 2846 | 157 | 5130 | 229 |
| 1995 | 160 | 1043 | 3107 | 1493 | 169 | 95 | 1786 | 150 | 6248 | 262 |
| 1996 | 155 | 1326 | 6106 | 1790 | 128 | 85 | 1371 | 113 | 3348 | 186 |
| 1997 | 141 | 1262 | 62137 | 2072 | 119 | 73 | 1055 | 106 | 2747 | 195 |
| 1998 | 182 | 1215 | 5111 | 2820 | 121 | 88 | 1119 | 82 | 2465 | 108 |
| 1999 | 237 | 1304 | 4149 | 3793 | 148 | 127 | 1189 | 74 | 2966 | 137 |
| 2000 | 405 | 1395 | 565 | 1964 | 53 | 119 | 1719 | 104 | 5482 | 103 |
| 2001 | 451 | 2485 | 55114 | 2935 | 58 | 140 | 2027 | 92 | 3309 | 122 |
| 2002 | 210 | 1286 | 6 146 | 3031 | 75 | 220 | 3800 | 206 | 6680 | 201 |
| 2003 | 151 | 1009 | 990 | 3108 | 65 | 171 | 1720 | 224 | 5899 | 146 |
| 2004 | 127 | 906 | 666 | 1980 | 72 | 83 | 974 | 150 | 3567 | 207 |
| 2005 | 87 | 380 | 025 | 921 | 59 | 73 | 768 | 33 | 1126 | 172 |
| 2006 | 54 | 359 | 967 | 989 | 119 | 56 | 598 | 47 | 1197 | 199 |
| 2007 | 94 | 713 | 31 | 1088 | 166 | 90 | 753 | 40 | 1811 | 239 |
| 2008 | 37 | 552 | 28 | 1325 | 163 | 100 | 1444 | 63 | 3361 | 318 |
| 2009 | 49 | 304 | 418 | 915 | 147 | 116 | 1571 | 100 | 3247 | 311 |
| 2010 | 34 | 418 | 840 | 970 | 183 | 63 | 1214 | 66 | 2350 | 302 |
| 2011 | 46 | 1091 | 155 | 2250 | 143 | 34 | 793 | 41 | 1433 | 324 |
| 2012 | 89 | 1295 | 5100 | 2215 | 185 | 35 | 909 | 56 | 2809 | 407 |
| 2013 | 41 | 896 | $6 \quad 42$ | 1236 | 191 | 42 | 1123 | 49 | 2342 | 405 |
| 2014 | 67 | 1247 | $7 \quad 73$ | 1889 | 236 | 60 | 1161 | 71 | 2781 | 647 |
| 2015 | 72 | 1183 | 379 | 3055 | 199 | 48 | 776 | 67 | 3985 | 338 |
| 2016 | 69 | 1151 | 110 | 1236 | 210 | 59 | 1165 | 83 | 1974 | 252 |
| 2017 | 28 | 303 | 3171 | 2225 | 158 | 0 | 0 | 41 | 727 | 74 |
| 2018 | 103 | 1478 | 8123 | 2166 | 267 | 55 | 694 | 55 | 1763 | 132 |
| 2019 | 99 | 1815 | 5103 | 3083 | 259 | 57 | 783 | 92 | 1929 | 120 |
| 2020 | 95 | 943 | 395 | 2425 | 306 | 61 | 572 | 61 | 1872 | 163 |



Table 4. Bss.27.4bc7a d-h: Sampling of commercial fishery seabass landings for length and age in France for lines, nets (A), Danish seines and other gears (B), and pelagic trawls and bottom trawls (C) (2017 real sampling excluding simulated). Nsamp is the number of landings (trips) sampled; Nfish is the number of fish measured.

| A. | FR_lines | FR_nets |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age |  | Length |  | Landings | Age |  | Length |  | Landings |
| Year | Nsamp | Nfish | Nsamp | Nfish |  | Nsamp | Nfish | Nsamp | Nfish |  |
| 2000 | NA | NA | 53 | 1613 | 305 | NA | NA | 2 | 72 | 108 |
| 2001 | NA | NA | 101 | 2659 | 375 | NA | NA | 1 | 5 | 110 |
| 2002 | NA | NA | 79 | 2076 | 349 | NA | NA | 0 | 0 | 128 |
| 2003 | NA | NA | 78 | 1732 | 438 | NA | NA | 1 | 4 | 152 |
| 2004 | NA | NA | 78 | 1748 | 381 | NA | NA | 6 | 84 | 150 |
| 2005 | NA | NA | 34 | 949 | 439 | NA | NA | 4 | 110 | 148 |
| 2006 | NA | NA | 73 | 1719 | 554 | NA | NA | 11 | 291 | 140 |
| 2007 | NA | NA | 69 | 2235 | 560 | NA | NA | 28 | 641 | 158 |
| 2008 | NA | NA | 41 | 1280 | 425 | NA | NA | 25 | 496 | 128 |
| 2009 | 12 | 211 | 33 | 1339 | 251 |  |  | 25 | 159 | 94 |
| 2010 | 4 | 169 | 10 | 334 | 278 |  |  | 49 | 615 | 160 |
| 2011 | 39 | 443 | 17 | 540 | 359 |  |  | 156 | 278 | 129 |
| 2012 | 37 | 385 | 10 | 681 | 295 |  |  | 60 | 408 | 142 |
| 2013 | 6 | 174 | 16 | 309 | 291 | 3 | 130 | 26 | 512 | 126 |
| 2014 |  |  | 10 | 299 | 285 |  |  | 29 | 218 | 163 |
| 2015 | 23 | 70 | 16 | 326 | 210 |  |  | 35 | 242 | 109 |
| 2016 |  |  | 2 | 84 | 156 | 5 | 67 | 32 | 293 | 64 |
| 2017 |  |  | 9 | 219 | 166 |  |  | 18 | 151 | 35 |
| 2018 |  |  | 4 | 208 | 151 | 9 | 45 | 9 | 45 | 74 |
| 2019 |  |  | 0 | 0 | 139 |  |  | 0 | 0 | 70 |
| 2020 |  |  | 27 | 703 | 164 |  |  | 13 | 193 | 78 |
| 2021 |  |  | 34 | 759 | 162 |  |  | 38 | 662 | 75 |


| B. | FR_danish seine |  |  |  |  | FR_other gears |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age |  | Length |  | Landings | Age |  | Length |  | Landings |
| Year | Nsamp | Nfish | Nsamp | Nfish |  | Nsamp | Nfish | Nsamp | Nfish |  |
| 2000 | NA | NA | 0 | 0 | 0 | NA | NA | 0 | 0 | 20 |
| 2001 | NA | NA | 0 | 0 | 0 | NA | NA | 0 | 0 | 27 |
| 2002 | NA | NA | 0 | 0 | 0 | NA | NA | 0 | 0 | 22 |
| 2003 | NA | NA | 0 | 0 | 0 | NA | NA | 0 | 0 | 23 |
| 2004 | NA | NA | 0 | 0 | 0 | NA | NA | 0 | 0 | 17 |
| 2005 | NA | NA | 0 | 0 | 0 | NA | NA | 0 | 0 | 17 |
| 2006 | NA | NA | 0 | 0 | 0 | NA | NA | 0 | 0 | 35 |
| 2007 | NA | NA | 0 | 0 | 0 | NA | NA | 0 | 0 | 24 |
| 2008 | NA | NA | 0 | 0 | 0 | NA | NA | 0 | 0 | 40 |
| 2009 |  |  | 0 | 0 | 27 |  |  | 0 | 0 | 127 |
| 2010 |  |  | 0 | 0 | 61 |  |  | 2 | 2 | 90 |
| 2011 |  |  | 2 | 6 | 43 |  |  | 36 | 292 | 62 |
| 2012 | 16 | 153 | 6 | 370 | 112 |  |  | 7 | 154 | 91 |
| 2013 |  |  | 2 | 28 | 18 |  |  | 1 | 1 | 82 |
| 2014 |  |  | 12 | 23 | 9 |  |  | 1 | 1 | 25 |
| 2015 | 10 | 36 | 0 | 12 | 26 |  |  | 0 | 0 | 16 |
| 2016 |  |  | 28 | 78 | 20 |  |  | 0 | 0 | 20 |
| 2017 |  |  | 14 | 42 | 22 |  |  | 0 | 0 | 40 |
| 2018 |  |  | 0 | 0 | 9 |  |  | 0 | 0 | 16 |
| 2019 |  |  | 0 | 0 | 21 |  |  | 0 | 0 | 22 |
| 2020 |  |  | 2 | 77 | 11 |  |  | 0 | 0 | 20 |
| 2021 |  |  | 2 | 33 | 16 |  |  | 0 | 0 | 17 |


| C. | FR_pelag | trawl |  |  |  | FR_bottom | rawl |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age |  | Length |  | Landings | Age |  | Length |  | Landings |
| Year | Nsamp | Nfish | Nsamp | Nfish |  | Nsamp | Nfish | Nsamp | Nfish |  |
| 2000 | NA | NA | 2 | 629 | 681 | NA | NA | 2 | 196 | 692 |
| 2001 | NA | NA | 0 | 0 | 659 | NA | NA | 0 | 0 | 713 |
| 2002 | NA | NA | 3 | 680 | 415 | NA | NA | 4 | 710 | 911 |
| 2003 | NA | NA | 4 | 753 | 773 | NA | NA | 8 | 998 | 1087 |
| 2004 | NA | NA | 6 | 938 | 820 | NA | NA | 12 | 887 | 1236 |
| 2005 | NA | NA | 11 | 1239 | 1319 | NA | NA | 14 | 689 | 1239 |
| 2006 | NA | NA | 16 | 2597 | 1420 | NA | NA | 11 | 1240 | 1110 |
| 2007 | NA | NA | 8 | 1800 | 841 | NA | NA | 11 | 588 | 1187 |
| 2008 | NA | NA | 8 | 1065 | 1012 | NA | NA | 18 | 1927 | 1145 |
| 2009 | 13 | 299 | 55 | 899 | 1098 | 20 | 164 | 93 | 1468 | 1052 |
| 2010 | 14 | 741 | 28 | 1299 | 1828 | 37 | 201 | 64 | 626 | 819 |
| 2011 | 38 | 1591 | 30 | 2309 | 1142 | 61 | 525 | 151 | 1955 | 791 |
| 2012 | 33 | 1587 | 9 | 1649 | 1143 | 51 | 478 | 87 | 1204 | 824 |
| 2013 | 17 | 737 | 10 | 1253 | 1516 | 34 | 344 | 73 | 2060 | 737 |
| 2014 | 11 | 202 | 23 | 455 | 242 | 50 | 326 | 137 | 2139 | 571 |
| 2015 |  |  | 12 | 158 | 107 | 57 | 203 | 76 | 1628 | 642 |
| 2016 |  |  | 6 | 48 | 17 | 103 | 407 | 183 | 1396 | 271 |
| 2017 |  |  | 0 | 0 | 6 | 37 | 120 | 126 | 495 | 33 |
| 2018 |  |  | 0 | 0 | 1 | 23 | 265 | 31 | 163 | 63 |
| 2019 |  |  | 0 | 0 | 1 | 13 | 73 | 22 | 104 | 76 |
| 2020 |  |  | 0 | 0 | 2 |  |  | 30 | 572 | 133 |
| 2021 |  |  | 2 | 33 | 3 |  |  | 61 | 1789 | 140 |

## Table 5. Bss.27.4bc7a d-h: Numbers-at-length in French commercial all-gears fishery landings (input to assessment at lengths $14-94 \mathrm{~cm}$ with $<20$ and $>88$ size classes empty).

|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 120 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 717 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| 122 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| 124 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| 126 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 63 |  | 0 |  | 0 | 0 | 0 |
| 128 | 0 | 0 | 0 | 3455 | 0 | 0 | 0 | 0 | 0 | 292 | 0 | 0 | 1219 | 0 | 0 | 291 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| 130 | 0 | 0 | 1015 | 13054 | 14 | 0 | 15689 | 0 | 0 | 473 | 0 | 0 | 0 | 146 | 0 | 346 | 71 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| 132 | 0 | 0 | 0 | 58717 | 13057 | 9903 | 32459 | 181 | 8250 | 2239 | 9811 | 1976 | 1583 | 0 | 3076 | 2678 | 1481 |  | 0 |  | 0 |  | 0 | 0 | 80 |
| 134 | 9931 | 17962 | 12469 | 105655 | 78811 | 29872 | 179130 | 4715 | 28986 | 10714 | 28290 | 13885 | 6518 | 1504 | 3620 | 5102 | 1440 |  | 137 |  | 0 |  | 0 | 0 | 0 |
| 136 | 34932 | 19809 | 38249 | 125326 | 127801 | 97890 | 285704 | 39335 | 229758 | 124925 | 169311 | 57121 | 85760 | 29667 | 33532 | 44175 | 2814 |  | 2646 |  | 0 |  | 0 | 194 | 90 |
| 138 | 85866 | 68920 | 46427 | 180475 | 124051 | 128022 | 217657 | 102714 | 263071 | 211881 | 177571 | 87842 | 172510 | 88507 | 68262 | 75546 | 4340 |  | 2523 |  | 91 |  | 0 | 1030 | 763 |
| 140 | 126730 | 76594 | 62503 | 119495 | 227214 | 231750 | 178250 | 146272 | 266408 | 225545 | 182105 | 128838 | 140273 | 149070 | 74871 | 93273 | 7417 |  | 3572 |  | 814 |  | 0 | 6255 | 7417 |
| 142 | 102836 | 98008 | 82461 | 145456 | 282390 | 266905 | 196868 | 145122 | 237160 | 193030 | 283064 | 187586 | 147895 | 146130 | 82684 | 115713 | 24816 |  | 9257 |  | 2444 |  | 2034 | 16127 | 24659 |
| 144 | 80478 | 109595 | 91064 | 104545 | 243107 | 344681 | 289998 | 164011 | 270810 | 222613 | 251956 | 201447 | 162333 | 123170 | 51365 | 122460 | 20422 |  | 14861 |  | 2954 |  | 2198 | 17867 | 22303 |
| 146 | 93344 | 106857 | 86723 | 130023 | 188494 | 270532 | 285451 | 130859 | 228996 | 238849 | 230227 | 199487 | 180752 | 140677 | 61292 | 95208 | 22427 |  | 9603 |  | 4379 |  | 1948 | 12708 | 20722 |
| 148 | 80934 | 77694 | 62163 | 115806 | 126685 | 239265 | 263272 | 100043 | 142650 | 155222 | 188149 | 194697 | 158490 | 127136 | 39844 | 59668 | 20653 |  | 7367 |  | 2606 |  | 635 | 9921 | 13639 |
| 150 | 55399 | 57055 | 55905 | 91915 | 72581 | 169478 | 200874 | 99210 | 112385 | 159658 | 186310 | 145447 | 130759 | 116842 | 38109 | 51436 | 15619 |  | 6801 |  | 3549 |  | 1246 | 5488 | 11644 |
| 152 | 52948 | 51658 | 46180 | 93878 | 82331 | 115269 | 119836 | 75929 | 74336 | 114530 | 109212 | 124239 | 107214 | 99156 | 29929 | 37860 | 10415 |  | 4599 |  | 2861 |  | 345 | 3890 | 7315 |
| 154 | 42094 | 36737 | 35998 | 48742 | 50633 | 62106 | 99509 | 74405 | 66260 | 84649 | 120550 | 92526 | 90638 | 103818 | 39911 | 21406 | 16034 |  | 3586 |  | 2702 |  | 456 | 2456 | 5362 |
| 156 | 26460 | 35839 | 26001 | 60839 | 60284 | 67741 | 99674 | 55147 | 48853 | 96257 | 71590 | 72471 | 78934 | 89197 | 32298 | 20681 | 9753 |  | 1012 |  | 2538 |  | 415 | 1515 | 4242 |
| 158 | 27357 | 22762 | 19019 | 31614 | 31334 | 61132 | 54522 | 46087 | 39689 | 51578 | 62211 | 46869 | 54869 | 59004 | 30016 | 13591 | 12328 |  | 2519 |  | 3581 |  | 0 | 1226 | 3181 |
| 160 | 23581 | 25834 | 14210 | 33688 | 19126 | 43591 | 45908 | 28056 | 29840 | 36547 | 31544 | 31690 | 35387 | 65851 | 21467 | 11946 | 7678 |  | 913 |  | 2008 |  | 0 | 1064 | 1996 |
| 162 | 14295 | 18773 | 11129 | 30691 | 23996 | 35774 | 23763 | 23057 | 28335 | 57472 | 19076 | 19998 | 33085 | 64579 | 16797 | 11776 | 7506 |  | 1120 |  | 1669 |  | 373 | 636 | 1617 |
| 164 | 18044 | 13532 | 16771 | 18823 | 14799 | 25788 | 20607 | 18091 | 14420 | 24016 | 62005 | 17624 | 17714 | 53482 | 16261 | 9356 | 4348 |  | 1369 |  | 1641 |  | 0 | 641 | 1199 |
| 166 | 10773 | 11068 | 11011 | 13230 | 10650 | 12456 | 14969 | 8715 | 12694 | 21415 | 26388 | 14720 | 15170 | 37744 | 8387 | 6653 | 2634 |  | 510 |  | 778 |  | 124 | 333 | 695 |
| 168 | 9903 | 9120 | 5447 | 7960 | 8569 | 13360 | 13976 | 8793 | 9039 | 27466 | 9340 | 7906 | 9374 | 23884 | 5579 | 2485 | 4465 |  | 315 |  | 463 |  | 124 | 378 | 882 |
| 170 | 5709 | 11771 | 4795 | 5374 | 4880 | 8908 | 9653 | 4835 | 6821 | 20198 | 8541 | 6114 | 8114 | 32512 | 8995 | 1163 | 1353 |  | 345 |  | 255 |  | 0 | 133 | 468 |
| 172 | 5721 | 5733 | 4559 | 5617 | 2974 | 8053 | 4521 | 2707 | 4714 | 12083 | 29128 | 2082 | 4147 | 14996 | 3027 | 660 | 956 |  | 408 |  | 47 |  | 0 | 62 | 141 |
| 174 | 2345 | 5345 | 1825 | 3275 | 2675 | 9811 | 3424 | 1962 | 1623 | 7551 | 1884 | 1163 | 2313 | 9001 | 642 | 628 | 219 |  | 652 |  | 0 |  | 0 | 0 | 121 |


|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 176 | 2595 | 2782 | 1260 | 1356 | 2567 | 5020 | 2883 | 1010 | 1257 | 979 | 2114 | 1096 | 1540 | 2640 | 773 | 431 | 0 |  | 92 |  | 0 |  | 0 | 0 | 76 |
| 178 | 2102 | 1691 | 357 | 297 | 548 | 2378 | 731 | 399 | 534 | 1765 | 182 | 476 | 1134 | 2073 | 0 | 9 | 127 |  | 718 |  | 0 |  | 0 | 0 | 5 |
| 180 | 888 | 583 | 155 | 783 | 425 | 1365 | 201 | 158 | 261 | 264 | 5525 | 148 | 282 | 176 | 198 | 16 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| 182 | 1021 | 296 | 109 | 112 | 149 | 107 | 261 | 37 | 8 | 1004 | 6097 | 104 | 451 | 1566 | 0 | 278 | 0 |  | 92 |  | 0 |  | 0 | 0 | 3 |
| 184 | 548 | 204 | 0 | 148 | 295 | 0 | 30 | 59 | 0 | 0 | 863 | 0 | 29 | 0 | 0 | 0 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| 186 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 27 | 1115 | 0 | 0 | 301 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| 188 | 0 | 61 | 0 | 0 | 149 | 0 | 0 | 0 | 0 | 0 | 1207 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 |

## Table 6. Bss.27.4bc7ad-h: Numbers-at-age in French commercial fishery landings, 2000-2021, all gears combined (with <2 year old age classes empty).

|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| a2 | 0 | 0 | 0 | 2611 | 3 | 0 | 3138 | 0 | 1208 | 878 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 |
| аз | 0 | 2651 | 8114 | 10800 | 4 | 24195 | 74600 | 5307 | 79917 | 17293 | 6830 | 0 | 0 | 0 | 0 | 0 | 0 | 98 | 0 | 0 | 34 | 1505 |
| a4 | 9440 | 55640 | 73892 | 364427 | 80483 | 77794 | 131099 | 73224 | 175402 | 92501 | 64071 | 3481 | 265 | 0 | 2300 | 10607 | 1850 | 1065 | 735 | 2251 | 7217 | 13637 |
| a5 | 222655 | 47734 | 125531 | 241694 | 627951 | 253455 | 564668 | 135809 | 545960 | 77583 | 74708 | 109582 | 17233 | 6417 | 4138 | 28960 | 20182 | 1605 | 5490 | 2392 | 17734 | 30515 |
| a6 | 273687 | 298773 | 90294 | 318445 | 438799 | 735235 | 361515 | 460583 | 401231 | 49868 | 135505 | 152702 | 216615 | 16487 | 36582 | 39482 | 18999 | 6568 | 1582 | 2391 | 37642 | 23147 |
| a7 | 139562 | 211740 | 236147 | 96562 | 297961 | 352182 | 841651 | 124606 | 456312 | 17981 | 112579 | 142896 | 266517 | 133787 | 54484 | 30417 | 18589 | 4507 | 1571 | 112 | 3170 | 34777 |
| a8 | 79413 | 90962 | 86108 | 254050 | 65297 | 443765 | 146484 | 139879 | 143871 | 17887 | 117368 | 121070 | 240104 | 88986 | 43902 | 60049 | 20496 | 6102 | 776 | 1682 | 4038 | 11829 |
| a9 | 47258 | 44742 | 31151 | 114829 | 131612 | 39104 | 253945 | 79978 | 147881 | 0 | 57320 | 264916 | 222113 | 109206 | 41316 | 27137 | 14682 | 3870 | 2704 | 454 | 6130 | 9287 |
| a10 | 43924 | 21074 | 23025 | 57883 | 77533 | 161572 | 13655 | 69214 | 40719 | 28366 | 22351 | 105282 | 172833 | 143048 | 4541 | 17343 | 14879 | 2085 | 70 | 231 | 4343 | 0 |
| a11 | 49293 | 39908 | 17823 | 26223 | 25416 | 69617 | 132370 | 33191 | 57341 | 1248 | 10523 | 58721 | 82759 | 103915 | 1742 | 5353 | 10075 | 1160 | 0 | 386 | 211 | 2465 |
| a12 | 20207 | 36007 | 14760 | 19879 | 14848 | 26314 | 84910 | 65868 | 17882 | 0 | 10414 | 24328 | 35102 | 47660 | 202 | 3804 | 0 | 727 | 306 | 0 | 1403 | 0 |
| a13 | 10767 | 17787 | 15912 | 14232 | 14254 | 17996 | 22068 | 68599 | 35092 | 0 | 7096 | 18672 | 21967 | 18471 | 178 | 2650 | 1423 | 649 | 0 | 0 | 0 | 976 |
| a14 | 4925 | 4394 | 9752 | 18088 | 13528 | 19238 | 6648 | 11131 | 12669 | 0 | 7652 | 4666 | 8640 | 16817 | 121 | 4485 | 131 | 351 | 0 | 0 | 0 | 471 |
| a15 | 4927 | 6838 | 3743 | 6600 | 7628 | 17974 | 6999 | 9034 | 5518 | 0 | 213 | 3149 | 2570 | 1275 | 0 | 0 | 81 | 455 | 0 | 0 | 0 | 0 |
| a16 | 10901 | 8034 | 1553 | 4028 | 5270 | 22718 | 16069 | 5486 | 6091 | 0 | 2322 | 410 | 1374 | 4149 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 7. Bss.27.4bc7ad-h: Numbers-at-age in the UK (England and Wales) bottom trawl, nets (with <2 year old age classes empty).

| UK OTB NET | a2 | a3 | a4 | a5 | a6 | a7 | a8 | a9 | a10 | a11 | a12 | a13 | a14 | a15 | a16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 65 | 11844 | 30828 | 6121 | 9692 | 1240 | 3914 | 9713 | 2454 | 2581 | 1320 | 343 | 841 | 286 | 892 |
| 1986 | 0 | 15673 | 20303 | 18759 | 3453 | 7662 | 704 | 3197 | 10503 | 1833 | 1403 | 2889 | 1222 | 1688 | 3595 |
| 1987 | 0 | 439 | 30263 | 58458 | 13753 | 2095 | 2437 | 656 | 726 | 5731 | 2565 | 1889 | 761 | 817 | 2796 |
| 1988 | 0 | 1930 | 20862 | 54472 | 41710 | 12803 | 1721 | 2315 | 780 | 451 | 5503 | 2024 | 1312 | 801 | 2589 |
| 1989 | 33394 | 5411 | 1223 | 7659 | 43911 | 26891 | 9002 | 3076 | 2901 | 1878 | 2896 | 8914 | 1499 | 1286 | 3436 |
| 1990 | 0 | 3035 | 2503 | 3770 | 16047 | 31459 | 21020 | 5042 | 2186 | 1463 | 846 | 1100 | 4837 | 353 | 2703 |
| 1991 | 1533 | 6933 | 36938 | 2381 | 1283 | 6576 | 18064 | 16248 | 7033 | 589 | 2617 | 2321 | 480 | 6659 | 3674 |
| 1992 | 0 | 15982 | 55550 | 33557 | 1183 | 796 | 1956 | 4750 | 4762 | 1230 | 451 | 433 | 139 | 497 | 3202 |
| 1993 | 0 | 657 | 81429 | 65981 | 21858 | 1351 | 627 | 1796 | 4803 | 3920 | 1500 | 710 | 735 | 475 | 2347 |
| 1994 | 2 | 1328 | 30970 | 369416 | 41472 | 16079 | 1130 | 294 | 2282 | 5842 | 4387 | 1596 | 650 | 646 | 3717 |
| 1995 | 0 | 5599 | 37064 | 81529 | 334815 | 17932 | 6931 | 702 | 415 | 1046 | 3440 | 3215 | 1846 | 2699 | 2680 |
| 1996 | 191 | 11473 | 43831 | 31632 | 64618 | 173733 | 8235 | 3622 | 216 | 315 | 454 | 1881 | 1688 | 534 | 1784 |
| 1997 | 0 | 2490 | 8501 | 64000 | 45238 | 39229 | 145407 | 8105 | 4456 | 632 | 640 | 294 | 2689 | 1712 | 2235 |
| 1998 | 0 | 1103 | 44997 | 49461 | 69489 | 25366 | 15136 | 41057 | 2671 | 860 | 96 | 96 | 385 | 623 | 811 |
| 1999 | 241 | 82 | 80414 | 146338 | 43841 | 28582 | 9612 | 6192 | 18072 | 1112 | 729 | 40 | 270 | 97 | 830 |
| 2000 | 0 | 9528 | 2584 | 151515 | 72747 | 11772 | 11046 | 4992 | 4636 | 8323 | 818 | 184 | 14 | 55 | 643 |
| 2001 | 614 | 11085 | 92408 | 29064 | 105169 | 25329 | 7388 | 8742 | 5811 | 8136 | 7522 | 804 | 768 | 69 | 759 |
| 2002 | 338 | 11495 | 43605 | 240476 | 16779 | 67647 | 16021 | 7450 | 8022 | 2682 | 3842 | 10166 | 645 | 193 | 568 |
| 2003 | 0 | 5698 | 75254 | 70415 | 154267 | 8719 | 38901 | 14072 | 4789 | 3196 | 2260 | 1599 | 3937 | 937 | 756 |
| 2004 | 0 | 4406 | 38270 | 214112 | 76652 | 95133 | 2733 | 12227 | 4039 | 1583 | 994 | 802 | 263 | 1029 | 221 |
| 2005 | 0 | 18910 | 135210 | 89202 | 124422 | 33796 | 30175 | 3112 | 7357 | 1390 | 1123 | 363 | 173 | 650 | 842 |
| 2006 | 0 | 20497 | 141335 | 144890 | 54069 | 56281 | 17344 | 24148 | 2207 | 3475 | 2277 | 859 | 210 | 188 | 1433 |
| 2007 | 0 | 955 | 33606 | 169272 | 96625 | 44423 | 34061 | 12877 | 14366 | 11530 | 4527 | 1621 | 11 | 254 | 428 |
| 2008 | 0 | 9338 | 110875 | 296983 | 139083 | 47617 | 19838 | 17332 | 8660 | 6128 | 852 | 793 | 988 | 317 | 824 |
| 2009 | 0 | 2659 | 73056 | 169969 | 172602 | 64997 | 19002 | 14443 | 9064 | 8631 | 3610 | 2235 | 1302 | 0 | 249 |
| 2010 | 0 | 319 | 77100 | 155258 | 118179 | 78410 | 28938 | 11821 | 6979 | 6043 | 2645 | 2083 | 2273 | 534 | 1663 |
| 2011 | 0 | 845 | 28630 | 124625 | 92582 | 71094 | 54338 | 31775 | 10438 | 11227 | 6347 | 2933 | 2203 | 675 | 1692 |
| 2012 | 0 | 1620 | 14135 | 166965 | 219883 | 61319 | 39609 | 31669 | 15268 | 9427 | 4092 | 3864 | 2546 | 538 | 930 |


| UK OTB NET | a2 |  | a3 | a4 | a5 | a6 | a7 | a8 | a9 | a10 | a11 | a12 | a13 | a14 | a15 | a16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 |  | 0 | 0 | 45016 | 60547 | 182858 | 117821 | 33448 | 30222 | 22727 | 17473 | 11825 | 2908 | 2687 | 2429 | 2133 |
| 2014 |  | 0 | 6622 | 31923 | 107001 | 58412 | 114826 | 78809 | 38859 | 27037 | 30548 | 19853 | 5152 | 1776 | 1857 | 1487 |
| 2015 |  | 0 | 50 | 3716 | 20172 | 45807 | 36830 | 63272 | 35025 | 17302 | 12685 | 10431 | 2917 | 7265 | 7308 | 966 |
| 2016 |  | 0 | 0 | 1591 | 7863 | 13991 | 31088 | 24925 | 40386 | 24807 | 10618 | 8218 | 4788 | 1960 | 2098 | 1528 |
| 2017 |  | 0 | 0 | 39 | 454 | 2176 | 1179 | 881 | 928 | 852 | 713 | 107 | 257 | 41 | 144 | 236 |
| 2018 |  | 0 | 130 | 4361 | 18582 | 26874 | 18792 | 9488 | 6826 | 4615 | 6186 | 5377 | 1562 | 1164 | 960 | 766 |
| 2019 |  | 0 | 105 | 2168 | 26492 | 29521 | 14508 | 9155 | 4501 | 4944 | 4192 | 4556 | 2635 | 1331 | 803 | 2067 |
| 2020 |  | 0 | 1058 | 4481 | 16161 | 85080 | 29885 | 12476 | 5890 | 3316 | 3182 | 2712 | 2768 | 2351 | 1456 | 1772 |
| 2021 |  | 0 | 270 | 2643 | 9490 | 20563 | 70224 | 23930 | 11483 | 5748 | 3157 | 2275 | 4297 | 2012 | 421 | 716 |

Table 8. Bss.27.4bc7ad-h: Numbers-at-age in the UK (England and Wales) lines (with <2 year old age classes empty).

|  | a2 | a3 | a4 | a5 | a6 | a7 | a8 | a9 | a10 | a11 | a12 | a13 | a14 | a15 | a16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0 | 9225 | 11491 | 3441 | 5902 | 891 | 1113 | 5133 | 1176 | 694 | 913 | 46 | 122 | 134 | 936 |
| 1986 | 0 | 577 | 8939 | 3343 | 933 | 2354 | 358 | 758 | 5428 | 960 | 871 | 953 | 573 | 645 | 1307 |
| 1987 | 0 | 108 | 1052 | 3719 | 2132 | 581 | 477 | 432 | 523 | 1578 | 845 | 211 | 167 | 179 | 1187 |
| 1988 | 0 | 33 | 1751 | 13389 | 5067 | 2398 | 551 | 1014 | 209 | 456 | 1863 | 895 | 715 | 523 | 977 |
| 1989 | 22 | 0 | 538 | 8171 | 36046 | 1842 | 371 | 104 | 208 | 58 | 215 | 1040 | 115 | 87 | 334 |
| 1990 | 0 | 305 | 82 | 185 | 1284 | 3456 | 2407 | 897 | 357 | 369 | 193 | 242 | 1261 | 81 | 828 |
| 1991 | 0 | 131 | 8420 | 471 | 177 | 792 | 4927 | 4024 | 1842 | 89 | 1229 | 1685 | 367 | 4831 | 2887 |
| 1992 | 0 | 1195 | 5473 | 5267 | 294 | 269 | 518 | 1193 | 1633 | 563 | 130 | 195 | 169 | 143 | 1411 |
| 1993 | 16 | 526 | 11652 | 11776 | 7569 | 590 | 289 | 931 | 3941 | 3344 | 1367 | 663 | 703 | 643 | 3789 |
| 1994 | 0 | 71 | 4059 | 119784 | 18540 | 9393 | 943 | 173 | 1754 | 5414 | 5570 | 1205 | 639 | 274 | 2790 |
| 1995 | 0 | 486 | 6943 | 21979 | 97509 | 7380 | 5313 | 480 | 699 | 831 | 5684 | 3696 | 1936 | 840 | 4733 |
| 1996 | 0 | 210 | 8804 | 12487 | 15338 | 57127 | 4566 | 4979 | 127 | 510 | 364 | 2521 | 1573 | 1300 | 2346 |
| 1997 | 59 | 454 | 3102 | 15613 | 11415 | 8287 | 50819 | 2853 | 1635 | 557 | 354 | 243 | 2195 | 1065 | 1570 |
| 1998 | 0 | 3676 | 8366 | 10920 | 22630 | 10485 | 6452 | 28231 | 2949 | 1091 | 138 | 196 | 793 | 1381 | 1254 |
| 1999 | 479 | 255 | 25158 | 37306 | 13589 | 13697 | 5288 | 5001 | 20522 | 1669 | 2038 | 247 | 777 | 315 | 3314 |
| 2000 | 0 | 421 | 294 | 19380 | 12402 | 2696 | 3285 | 1476 | 1248 | 4697 | 330 | 258 | 16 | 88 | 559 |
| 2001 | 54 | 471 | 7385 | 1392 | 17864 | 7702 | 2027 | 3239 | 1685 | 1761 | 3774 | 440 | 301 | 27 | 420 |
| 2002 | 30 | 729 | 2609 | 14173 | 2686 | 17358 | 7757 | 2621 | 5179 | 1463 | 1766 | 3687 | 322 | 101 | 180 |
| 2003 | 0 | 80 | 7166 | 7917 | 25014 | 2167 | 10164 | 3262 | 1473 | 982 | 796 | 681 | 1704 | 186 | 166 |
| 2004 | 0 | 279 | 1697 | 13884 | 8601 | 17310 | 2398 | 6365 | 3626 | 1181 | 1189 | 1172 | 406 | 2243 | 143 |
| 2005 | 0 | 621 | 2669 | 5059 | 14699 | 5529 | 6985 | 589 | 5697 | 1845 | 236 | 1307 | 33 | 189 | 606 |
| 2006 | 0 | 44 | 16121 | 35990 | 13714 | 22306 | 5794 | 12717 | 1644 | 3135 | 1258 | 305 | 358 | 1016 | 734 |
| 2007 | 0 | 22 | 6611 | 31578 | 28396 | 14511 | 17834 | 8499 | 10951 | 5163 | 3121 | 5119 | 85 | 344 | 485 |
| 2008 | 0 | 199 | 5010 | 27319 | 42071 | 21561 | 12265 | 12566 | 5458 | 4960 | 1372 | 1032 | 3431 | 198 | 992 |
| 2009 | 0 | 315 | 8415 | 19843 | 33661 | 25695 | 12017 | 9320 | 5021 | 5371 | 4748 | 811 | 1075 | 0 | 0 |
| 2010 | 0 | 814 | 7029 | 45515 | 54766 | 39716 | 15835 | 5147 | 2395 | 2910 | 706 | 522 | 359 | 81 | 277 |
| 2011 | 0 | 8 | 5209 | 11538 | 24667 | 19293 | 16668 | 13032 | 4947 | 6066 | 2695 | 1941 | 2187 | 522 | 657 |
| 2012 | 0 | 91 | 1695 | 18362 | 28593 | 23507 | 22946 | 17909 | 10199 | 7725 | 2994 | 2672 | 2158 | 596 | 820 |
| 2013 | 0 | 0 | 1187 | 6979 | 35135 | 32251 | 18057 | 14762 | 10333 | 10543 | 6106 | 3730 | 2886 | 1957 | 1938 |
| 2014 | 0 | 980 | 4985 | 26081 | 20743 | 39548 | 28357 | 15323 | 12440 | 12413 | 8018 | 4889 | 1976 | 1673 | 1322 |
| 2015 | 0 | 6 | 1834 | 5941 | 23369 | 22221 | 31442 | 19014 | 10344 | 8210 | 7036 | 2504 | 3136 | 744 | 798 |
| 2016 | 0 | 0 | 742 | 7020 | 11858 | 20142 | 15479 | 25838 | 13362 | 7406 | 5904 | 4674 | 2548 | 3894 | 2567 |
| 2017 | 0 | 0 | 1734 | 4007 | 5766 | 2324 | 2362 | 1036 | 4159 | 993 | 356 | 469 | 202 | 475 | 330 |
| 2018 | 0 | 454 | 6992 | 23652 | 41538 | 31173 | 17352 | 16753 | 11214 | 14117 | 9044 | 4650 | 3791 | 2220 | 3945 |
| 2019 | 0 | 85 | 3010 | 36477 | 41315 | 26099 | 16791 | 9320 | 10364 | 11061 | 9434 | 5936 | 3248 | 2068 | 4291 |
| 2020 | 0 | 431 | 3437 | 11667 | 90256 | 53606 | 27720 | 13526 | 7890 | 7117 | 5823 | 5194 | 3678 | 2127 | 3185 |
| 2021 | 0 | 206 | 5501 | 21294 | 39528 | 93379 | 41709 | 26913 | 12827 | 6814 | 10191 | 4230 | 3570 | 3866 | 5300 |

Table 9. Bss.27.4bc7ad-h: Numbers-at-age in the UK (England and Wales) midwater pair trawl fleet (no samples for 1997, 2013-2017, 2019-present) (with <3 year old age classes empty).

|  | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 | Age11 | Age12 | Age13 | Age14 | Age15 | Age16+ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 0 | 289 | 796 | 3892 | 71666 | 5583 | 1648 | 21 | 334 | 154 | 622 | 485 | 199 | 559 |
| 1998 | 0 | 245 | 5979 | 11845 | 8553 | 8135 | 25138 | 2517 | 345 | 93 | 53 | 119 | 893 | 569 |
| 1999 | 0 | 2983 | 18409 | 15106 | 27147 | 13818 | 18060 | 43097 | 4389 | 1686 | 324 | 387 | 308 | 2689 |
| 2000 | 15 | 60 | 2476 | 7587 | 3270 | 4497 | 1459 | 2830 | 7077 | 634 | 174 | 39 | 96 |  |
| 2001 | 0 | 179 | 899 | 19777 | 20290 | 7042 | 5268 | 3124 | 2845 | 9666 | 857 | 636 | 123 | 261 |
| 2002 | 3 | 37 | 2380 | 1578 | 24087 | 9693 | 6297 | 5978 | 450 | 5664 | 9215 | 0 | 0 | 530 |
| 2003 | 0 | 2689 | 10619 | 39257 | 7971 | 40551 | 10293 | 3162 | 3254 | 618 | 169 | 4043 | 77 | 281 |
| 2004 | 7 | 1254 | 12502 | 14372 | 48109 | 3199 | 20694 | 8010 | 353 | 1797 | 1141 | 91 | 968 | 18 |
| 2005 | 0 | 114 | 2103 | 15321 | 14397 | 17408 | 1907 | 5182 | 0 | 1831 | 99 | 0 | 40 | 599 |
| 2006 | 0 | 227 | 567 | 608 | 4076 | 1423 | 3085 | 254 | 176 | 111 | 0 | 0 | 0 | 53 |
| 2007 | 0 | 385 | 2517 | 7038 | 5387 | 6833 | 2795 | 1900 | 631 | 807 | 12 | 37 | 19 | 121 |
| 2008 | 45 | 445 | 1540 | 3279 | 1787 | 1412 | 1557 | 755 | 960 | 30 | 183 | 490 | 0 | 40 |
| 2009 | 0 | 90 | 635 | 2175 | 2596 | 843 | 784 | 168 | 298 | 173 | 11 | 169 | 0 | 0 |
| 2010 | 9 | 36 | 1741 | 5546 | 8261 | 6678 | 4755 | 403 | 3786 | 152 | 294 | 313 | 551 | 50 |
| 2011 | 0 | 255 | 4397 | 10231 | 13640 | 15909 | 13642 | 4424 | 4233 | 2773 | 1688 | 1003 | 264 | 423 |
| 2012 | 0 | 391 | 4461 | 10776 | 10016 | 8757 | 5789 | 2741 | 1134 | 290 | 433 | 143 | 127 | 226 |
| 2015 | 0 | 7 | 23 | 85 | 103 | 137 | 30 | 6 | 3 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 2 | 9 | 5 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

## Table 10. Bss.27.4bc7ad-h: Numbers of trips sampled for discards by Cefas (UK): 2002-2021, by gear group and area

| Division \& fleet |  | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (a) bottom otter trawls |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 4 | 16 | 34 | 56 | 37 | 41 | 85 | 58 | 49 | 46 | 42 | 54 | 30 | 53 | 45 | 12 | 0 | 1 | 3 |  |  |
| 7.afg |  | 8 | 15 | 23 | 8 | 11 | 43 | 50 | 28 | 22 | 22 | 22 | 12 | 14 | 16 | 2 | 0 | 0 | 5 |  |  |
| 7.d |  | 1 | 2 | 4 | 3 | 1 | 2 | 1 | 6 | 7 | 9 | 4 | 5 | 7 | 3 | 13 | 1 | 1 | 7 |  | 11 |
| 7.eh |  | 9 | 24 | 37 | 31 | 49 | 90 | 87 | 38 | 29 | 32 | 29 | 45 | 73 | 68 | 29 | 0 | 10 | 18 | 4 | 9 |
| Total |  | 34 | 75 | 120 | 79 | 102 | 220 | 196 | 121 | 104 | 105 | 109 | 92 | 147 | 132 | 56 | 1 | 12 | 33 | 4 | 20 |

(b) Fixed/driftnets

|  | 4 | 0 | 0 | 2 | 1 | 11 | 31 | 15 | 20 | 15 | 11 | 13 | 18 | 10 | 7 | 0 | 0 | 0 | 0 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7.afg |  | 3 | 7 | 5 | 3 | 7 | 8 | 9 | 10 | 7 | 16 | 22 | 16 | 25 | 12 | 3 | 0 | 0 | 0 |
| 7.d |  | 0 | 0 | 1 | 0 | 0 | 17 | 6 | 4 | 1 | 7 | 10 | 42 | 25 | 17 | 10 | 0 | 0 | 16 |
| 7.eh | 1 | 5 | 9 | 2 | 3 | 16 | 10 | 14 | 19 | 17 | 25 | 24 | 24 | 15 | 0 | 0 | 0 | 0 |  |
| Total | 4 | 12 | 17 | 6 | 21 | 72 | 40 | 48 | 42 | 51 | 70 | 100 | 84 | 51 | 13 | 0 | 0 | 16 |  |
| (c) Lines |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0 | 1 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 |  |
| 7.afg | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.d |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 33 | 2 | 0 | 0 |
| 7.eh | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 8 | 5 | 4 | 0 | 0 | 0 |  |
| Total | 0 | 1 | 1 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 2 | 1 | 10 | 6 | 37 | 2 | 0 | 0 |  |

(d) Midwater trawls

|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 7.dg | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.eh | 0 | 1 | 1 | 1 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 |  |
| 7.dal | 1 | 1 | 1 | 3 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 2 |  | 0 | 0 | 0 | 0 | 0 |  |

(e) Other gears

| Division \& fleet | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 20202021 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7.afg | 4 | 11 | 8 | 4 | 9 | 1 | 2 | 3 | 3 | 1 | 4 | 8 |  | 0 | 5 | 0 | 0 | 0 |  |
| 7.d | 0 | 1 | 5 | 2 | 3 | 1 | 1 | 2 | 4 | 1 | 2 | 3 | 1 | 2 | 0 | 0 | 0 | 0 |  |
| 7.eh | 10 | 17 | 27 | 16 | 24 | 32 | 18 | 13 | 17 | 27 | 22 | 21 | 14 | 15 | 1 | 0 | 0 | 0 |  |
| Total | 22 | 34 | 50 | 23 | 38 | 35 | 22 | 25 | 30 | 37 | 32 | 42 | 15 | 17 | 12 | 0 | 0 | 0 |  |

Table 11. Bss.27.4bc7ad-h: Estimated annual numbers and weight of seabass retained and discarded by UK using fixed or driftnets, otter trawl, beam trawl and lines fleets in areas 4, 7.d, 7.eh and 7.afg, based on at-sea sampling, and raised from landings in sampled strata to landings in all strata. Numbers of sampled trips (Ntrip) are shown.

|  | Otter trawl |  |  |  | Nets |  |  |  | Beam trawl |  |  |  | Lines |  |  |  | Total OTB, nets, lines and BTS |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | discards | retained | rate (\%) | Ntirip | discards | retained | rate \% | Ntirip | discards | retained | rate \% | Ntirip | discards | retained | rate (\%) | Ntirip | discards | retained | rate\% |
| 2002 | 17 | 161 | 9 | 34 | 0 | 201 | 0 | 4 | 0.2 | 24 | 0.7 | - | - | - | - | - | 17 | 386 | 4 |
| 2003 | 16 | 207 | 7 | 75 | 0 | 146 | 0 | 12 | 1.9 | 21 | 8.1 | - | - | - | - | - | 18 | 374 | 5 |
| 2004 | 59 | 173 | 25 | 120 | 0 | 207 | 0 | 17 | 0.3 | 24 | 1.3 | - | - | - | - | - | 59 | 404 | 13 |
| 2005 | 6 | 181 | 3 | 79 | 90 | 172 | 34 | 6 | 2.4 | 15 | 13.7 | - | - | - | - | - | 99 | 368 | 21 |
| 2006 | 34 | 160 | 17 | 102 | 19 | 199 | 9 | 21 | 0.4 | 14 | 2.5 | - | - | - | - | - | 53 | 373 | 12 |
| 2007 | 49 | 173 | 22 | 220 | 1 | 239 | 0.4 | 72 | 0.0 | 19 | 0.0 | - | - | - | - | - | 50 | 432 | 10 |
| 2008 | 5 | 196 | 3 | 196 | 3 | 318 | 0.9 | 40 | 1.2 | 21 | 5.6 | - | - | - | - | - | 9 | 535 | 2 |
| 2009 | 85 | 175 | 33 | 121 | 0 | 311 | 0.1 | 48 | 0.2 | 10 | 1.5 | - | - | - | - | - | 86 | 495 | 15 |
| 2010 | 49 | 150 | 25 | 104 | 1 | 302 | 0.3 | 42 | 1.2 | 6 | 17.1 | - | - | - | - | - | 51 | 458 | 10 |
| 2011 | 8 | 137 | 6 | 105 | 14 | 324 | 4.2 | 51 | 0.0 | 5 | 0.0 | - | - | - | - | - | 22 | 467 | 5 |
| 2012 | 27 | 157 | 15 | 109 | 2 | 407 | 0.5 | 70 | 0.0 | 5 | 0.0 | - | - | - | - | - | 29 | 569 | 5 |
| 2013 | 4 | 125 | 3 | 92 | 2 | 405 | 0.4 | 100 | 1.1 | 4 | 20.1 | - | - | - | - | - | 6 | 534 | 1 |
| 2014 | 1 | 104 | 1 | 147 | 6 | 647 | 0.9 | 84 | 0.0 | 8 | 0.0 | - | - | - | - | - | 7 | 758 | 1 |
| 2015 | 6 | 77 | 7 | 132 | 1 | 340 | 0.4 | 51 | 0.0 | 8 | 0.0 | - | - | - | - | - | 7 | 425 | 2 |
| 2016 | 35 | 52 | 40 | 56 | 8 | 252 | 3 | 13 | 0.1 | 23 | 0.0 |  | 8.4 | 210.0 | 4.0 | 37.0 | 52 | 537 | 9 |
| 2017* | 0 | 35 | 1 | 1 | - | 74 | - | 0 | - | 16 | - | 0 | 11 | 147 | 7 | 2 |  | 272 | - |
| 2018* | 11 | 13 | 46 | 5 | - | 132 | - | 0 | 15 | 13 | 54 | 7 | - | 267 | - | 0 | 26 | 425 | 6 |
| 2019* | 83 | 15 | 85 | 3 | 6 | 120 | 5 |  | 0.1 | 12 | 1 |  | 0 | 258 | 0 |  | 89 | 410 | 18 |
| 2020* | 12 | 27 | 44 | 1 | 0 | 163 | 0 | 3 | 13 | 15 | 87 | 0 | 0 | 306 | 0 | 0 | 25 | 522 | 5 |
| 2021* | 8 | 33 | 19 | 20 | 0 | 195 | 0 | 0 | 0 | 13 | 0 | 0 | 0 | 361 | 0 | 20 | 8 | 607 | 1 |

[^16]Table 12. Bss.27.4bc7ad-h: Number of fishing trips sampled for retained and discarded weight of seabass on French vessels using different gear types: 2009-2020. (Data are clearly underestimated from 2015 and are not used in assessment).

| Pelagic trawl FR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 773 | 0.00\% | NA |  |  |
| 2004 | 0 | 820 | 0.00\% | NA |  |  |
| 2005 | 0 | 1319 | 0.00\% | NA |  |  |
| 2006 | 0 | 1420 | 0.00\% | NA |  |  |
| 2007 | 0 | 841 | 0.00\% | NA | 12 | 2 |
| 2008 | 2 | 1012 | 0.20\% | 3.93 | 21 | 4 |
| 2009 | 21.2 | 1098 | 1.89\% | 0.05 |  |  |
| 2010 | 7.4 | 1828 | 0.40\% | 0.71 | 35 | 106 |
| 2011 | 7.2 | 1142 | 0.63\% | 0.12 | 9 | 46 |
| 2012 | 0.9 | 1143 | 0.08\% | 2.38 | 7 | 29 |
| 2013 | 0.3 | 1516 | 0.02\% | 2 |  |  |
| 2014 | 0 | 242 | 0.00\% | NA |  |  |
| 2015 | 11.7 | 107 | 9.86\% | 0.03 | 32 | 5 |
| 2016* | 0.5 | 17.43081 | 2.79\% | NA | 19 | 2 |
| 2017* |  | 6 |  | NA | 0 | 0 |
| 2018* | 0.2 | 1 | 17\% |  | 28 | 1 |
| 2019* |  | 1 |  |  |  |  |
| 2020* | - | 2 | - | - | 0 | 0 |
| 2021* | 0 | 3 |  |  |  |  |
| bottom trawlFR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| 2003 | 73.8 | 1087 | 6.36\% | 0.35 | 18 | 26 |
| 2004 |  | 1236 | NA | NA | 24 | 3 |
| 2005 | 43.9 | 1239 | 3.42\% | 0.9 |  |  |
| 2006 | 42.9 | 1110 | 3.72\% | 1.07 | 24 | 36 |
| 2007 | 9.6 | 1187 | 0.80\% | 0.73 |  |  |
| 2008 | 40.7 | 1145 | 3.43\% | 0.94 | 57 | 63 |
| 2009 |  | 1052 | NA | NA | 143 | 102 |
| 2010 | 76.6 | 819 | 8.55\% | 0.32 | 137 | 5 |
| 2011 | 27.2 | 791 | 3.32\% | 0.46 | 122 | 57 |
| 2012 | 24.5 | 824 | 2.89\% | 0.23 | 151 | 118 |
| 2013 | 26.3 | 737 | 3.45\% | 0.37 | 139 | 145 |
| 2014 |  | 571 | NA | NA | 133 | 29 |
| 2015 | 35.4 | 642 | 5.23\% | 0.49 | 189 | 356 |
| 2016* | 126.9 | 271 | 31.86\% | NA | 512 | 90 |
| 2017* | 156 | 178 | 47\% | NA | 61 | 141 |
| 2018* | 32 | 72 | 31\% |  | 217 | 71 |
| 2019* | 76 | 76 | 50\% |  | 9 | 31 |
| 2020* | 3 | 133 | 2\% | - | 4 | 4 |


| 2021* | 11 | 140 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| netsFR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| 2003 | 31.7 | 152 | 17.26\% | 1.2 |  |  |
| 2004 | 77.6 | 150 | 34.09\% | 0.1 |  |  |
| 2005 | 0 | 148 | 0.00\% | NA |  |  |
| 2006 | 125.5 | 140 | 47.27\% | 0.34 |  |  |
| 2007 | 2.2 | 158 | 1.37\% | 0.61 | 32 | 2 |
| 2008 | 0.5 | 128 | 0.39\% | 0.79 |  |  |
| 2009 | 6.4 | 94 | 6.37\% | 0.41 | 196 | 3 |
| 2010 | 6.1 | 160 | 3.67\% | 0.29 | 108 | 5 |
| 2011 | 9 | 129 | 6.52\% | 0.35 |  |  |
| 2012 | 11.8 | 142 | 7.67\% | 0.55 | 269 | 9 |
| 2013 | 21.6 | 126 | 14.63\% | 0.18 | 173 | 2 |
| 2014 | 21.7 | 163 | 11.75\% | 0.11 | 118 | 3 |
| 2015 | 14.7 | 109 | 11.88\% | 0.2 | 217 | 8 |
| 2016* | 19.4 | 64 | 23.25\% | NA | 258 | 209 |
| 2017* | 0.7 | 34 | 2\% | NA | 0 | 0 |
| 2018* | 2 | 74 | 3\% |  | 101 | 17 |
| 2019* | 3 | 70 | 4\% |  |  |  |
| 2020* | - | 78 | - | - | 12 | 0 |
| 2021* | - | 75 | - | - |  |  |
| linesFR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| 2003 | 0 | 438 | 0.00\% | NA |  |  |
| 2004 | 0 | 381 | 0.00\% | NA |  |  |
| 2005 | 0 | 439 | 0.00\% | NA |  |  |
| 2006 | 0 | 554 | 0.00\% | NA |  |  |
| 2007 | 0 | 560 | 0.00\% | NA |  |  |
| 2008 | 100.3 | 425 | 19.09\% | 0.35 |  |  |
| 2009 | 5.6 | 251 | 2.18\% | 0.71 | 17 | 21 |
| 2010 | 3.9 | 278 | 1.38\% | 1.24 |  |  |
| 2011 | 13.1 | 359 | 3.52\% | 0.35 |  |  |
| 2012 | 15.8 | 295 | 5.08\% | 0.26 |  |  |
| 2013 | 14.2 | 291 | 4.65\% | 0.45 |  |  |
| 2014 | 15.8 | 285 | 5.25\% | 0.4 |  |  |
| 2015 | 7.4 | 210 | 3.40\% | 0.32 | 28 | 21 |
| 2016* |  | 156 |  | NA |  |  |
| 2017* |  | 166 |  | NA | 0 | 0 |
| 2018* |  | 151 |  |  | 0 | 0 |
| 2019* |  | 139 |  |  |  |  |
| 2020* | - | 164 | - | - | 0 | 0 |
| 2021* | 2 | 162 |  |  |  |  |


| OtherFR | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0 | 23 | 0.00\% | NA |  |  |
| 2004 | 6.6 | 17 | 27.97\% | NA |  |  |
| 2005 | 0 | 17 | 0.00\% | NA |  |  |
| 2006 | 0 | 35 | 0.00\% | NA |  |  |
| 2007 | 0 | 24 | 0.00\% | NA |  |  |
| 2008 | 0 | 40 | NA | NA |  |  |
| 2009 | 0 | 127 | NA | NA |  |  |
| 2010 | 0 | 90 | 0.00\% | NA |  |  |
| 2011 | 44.8 | 62 | 41.95\% | 5.97 |  |  |
| 2012 | 1.1 | 91 | 1.19\% | 0.25 | 6 | 9 |
| 2013 | 0 | 82 | 0.00\% | NA |  |  |
| 2014 | 0 | 25 | 0.00\% | NA | 130 | 96 |
| 2015 | 11 | 11 | 50.00\% | 0.58 |  |  |
| 2016* | 5.9 | 19.82406 | 22.94\% | NA | 64 | 9 |
| 2017* | 5 | 58 | 8\% | NA | 0 | 0 |
| 2018* |  | 15 |  |  | 0 | 0 |
| 2019* |  | 43 |  |  |  |  |
| 2020* | - | 20 | - | - | - | - |
| 2021* | - | 33 | - | - | - | - |
| FR_ALL | discards (t) | Landings (t) | discard rates | cv indicator | Nb trip sampled | Nb fish sampled |
| 2003 | 105.5 | 2473 | 4\% |  | 18 | 26 |
| 2004 | 84.2 | 2604 | 3\% |  | 24 | 3 |
| 2005 | 43.9 | 3162 | 1\% |  | 0 | 0 |
| 2006 | 168.4 | 3259 | 5\% |  | 24 | 36 |
| 2007 | 11.8 | 2770 | 0\% |  | 44 | 4 |
| 2008 | 143.5 | 2750 | 5\% |  | 78 | 67 |
| 2009^ | 33.2 | 2622 | 1\% |  | 356 | 126 |
| 2010^ | 94 | 3175 | 3\% |  | 280 | 116 |
| 2011^ | 101.3 | 2483 | 4\% | 7.25 | 131 | 103 |
| 2012^ | 54.1 | 2495 | 2\% | 3.67 | 433 | 165 |
| 2013^ | 62.4 | 2752 | 2\% |  | 312 | 147 |
| 2014^ | 37.5 | 1286 | 3\% |  | 381 | 128 |
| 2015^ | 80.2 | 1079 | 7\% | 1.62 | 466 | 390 |
| 2016* | 152.7 | 529 | 22\% |  | 853 | 310 |
| 2017* | 161.7 | 442 | 27\% |  | 61 | 141 |
| 2018* | 34.2 | 313 | 10\% |  |  |  |
| 2019* | 79.2 | 329 | 19\% |  | 9 | 31 |
| 2016** | 155.6 | 529 | 23\% |  |  |  |
| 2017** | 270.9 | 442 | 38\% |  |  |  |
| 2018** | 456.4 | 313 | 59\% |  |  |  |


| $2019^{* *}$ | 374.6 | 329 | $53 \%$ |
| :---: | :---: | :---: | :---: |
| $2020^{* *}$ | 313 | 409 | $43 \%$ |
| $2021^{* *}$ | 404 | 413 | $49 \%$ |

${ }^{\wedge}$ included in the assessment (source on-board sampling programme).

* not included in the assessment (source on-board sampling programme).
** included in the assessment (source logbook data).

Table 13. Bss.27.4bc7ad-h: Values of expected recreational F reductions associated with management measures applied to bss.27.4bc7ad-h since 2015. Frec multiplier represents the recreational F relative to 2012. Note that the emergency measures were implemented part way through $\mathbf{2 0 1 5}$, so the reduction was applied for half the year.

| Year | Management scenario |  |  | Frec Multiplier |
| :--- | :--- | :--- | :--- | :--- |
|  | MCRS | Bag limit | Open season |  |
| Pre-2015 | 36 cm | none | All year | 1.000 |
| 2015 Jan-Jun | 36 cm | none | All year | 0.821 |
| 2015 Jul-Dec | 42 cm | three fish | 3 months | 0.282 |
| $2016 \& 2017$ | 42 cm | one fish | 7 months | 0.191 |
| 2018 | 42 cm | one fish | 9 months | 0.312 |
| 2019 | 42 cm | two fish | 9 months | 0.464 |
| 2020 | 42 cm | two fish | 9 months | 0.464 |
| 2021 | 42 cm | two fish | 0.464 |  |
| 2021 |  |  |  |  |

Table 14. Bss.27.4bc7ad-h: Time-series of Cefas Solent autumn survey of juvenile seabass. Indices were revised in 2020 and updated in the assessment. A change in trawl design took place in 1993, and calibration factors are applied.

| Year | Solent Index prior to 2020 revision | 2021 Solent Index |
| :---: | :---: | :---: |
| 1986 | 5.84 | 5.84 |
| 1987 | 2.6 | 2.6 |
| 1989 | 7.05 | 7.05 |
| 1990 | 3.98 | 3.98 |
| 1991 | 3.32 | 3.32 |
| 1992 | 19.7 | 19.7 |
| 1993 | 14.63 | 14.63 |
| 1994 | 5.46 | 6.69 |
| 1995 | 10.24 | 10.53 |
| 1996 | 6.06 | 6.35 |
| 1997 | 38.2 | 40.4 |
| 1998 | 7.34 | 7.22 |
| 1999 | 20.91 | 19.02 |
| 2000 | 17.46 | 17.8 |
| 2001 | 39.91 | 42.69 |
| 2002 | 11.7 | 13.95 |
| 2003 | 13.55 | 14.18 |
| 2005 | 21.93 | 23.46 |
| 2006 | 19.73 | 19.76 |
| 2007 | 5.5 | 5.5 |
| 2008 | 25.52 | 25.52 |
| 2009 | 19.83 | 19.83 |
| 2011 | 4.05 | 4.05 |
| 2013 | 1.52 | 1.56 |
| 2014 | 1.4 | 1.45 |
| 2015 | 7.44 | 7.45 |
| 2016 | 6.03 | 6.2 |
| 2017 | 3.54 | 3.54 |
| 2018 | 2.66 | 2.66 |
| 2019 |  | 1.95 |
| 2020 |  | 4.92 |
| 2021 |  | 3.59 |

Table 15. Bss.27.4bc7ad-h: Numbers-at-age in Solent survey 1986-present: updated time-series of Cefas Solent autumn survey of juvenile seabass.

|  | Age 2 | Age 3 | Age 4 |
| :---: | :---: | :---: | :---: |
| 1986 | 0.27 | 4.26 | 1.31 |
| 1987 | 0.05 | 0.28 | 2.27 |
| 1989 | 6.68 | 0.37 | 0.00 |
| 1990 | 2.81 | 1.15 | 0.02 |
| 1991 | 3.08 | 0.21 | 0.03 |
| 1992 | 0.95 | 18.59 | 0.16 |
| 1993 | 6.65 | 3.59 | 4.39 |
| 1994 | 3.67 | 2.69 | 0.34 |
| 1995 | 4.19 | 5.88 | 0.46 |
| 1996 | 5.86 | 0.38 | 0.12 |
| 1997 | 33.78 | 6.54 | 0.08 |
| 1998 | 1.23 | 5.41 | 0.58 |
| 1999 | 17.62 | 0.59 | 0.82 |
| 2000 | 5.91 | 11.86 | 0.03 |
| 2001 | 36.70 | 4.21 | 1.77 |
| 2002 | 7.07 | 6.56 | 0.31 |
| 2003 | 8.51 | 5.07 | 0.60 |
| 2005 | 14.21 | 8.37 | 0.88 |
| 2006 | 9.53 | 9.21 | 1.02 |
| 2007 | 3.42 | 1.78 | 0.30 |
| 2008 | 18.52 | 6.66 | 0.34 |
| 2009 | 13.19 | 6.31 | 0.32 |
| 2011 | 2.25 | 1.39 | 0.41 |
| 2013 | 1.38 | 0.08 | 0.10 |
| 2014 | 0.76 | 0.67 | 0.02 |
| 2015 | 6.95 | 0.44 | 0.05 |
| 2016 | 3.86 | 2.24 | 0.11 |
| 2017 | 0.86 | 2.56 | 0.12 |
| 2018 | 2.17 | 0.32 | 0.18 |
| 2019 | 0.57 | 1.36 | 0.02 |
| 2020 | 3.85 | 0.87 | 0.20 |
| 2021 | 1.94 | 1.60 | 0.05 |

Table 16. Bss.27.4bc7ad-h: Seabass indices of abundance 2000-2014 (swept area) from the Channel Groundfish Survey. The relative standard error CV is the log-transformed value used in SS3 (sqrt(loge (1+CV^2)).

| year | Total <br> hauls | No. hauls with seabass | Percentage of hauls with seabass | Mean no. seabass per positive haul | Swept-area abundance index | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 68 | 6 | 9 | 2 | 245776 | 0.15 |
| 1989 | 61 | 3 | 5 | 1 | 77716 | 0.58 |
| 1990 | 75 | 8 | 11 | 8 | 1129914 | 0.12 |
| 1991 | 79 | 19 | 24 | 9 | 4250636 | 0.03 |
| 1992 | 60 | 23 | 38 | 13 | 2617986 | 0.11 |
| 1993 | 65 | 21 | 32 | 8 | 2299919 | 0.10 |
| 1994 | 86 | 19 | 22 | 5 | 1097828 | 0.11 |
| 1995 | 166 | 17 | 10 | 5 | 1021741 | 0.09 |
| 1996 | 134 | 26 | 19 | 3 | 1224238 | 0.13 |
| 1997 | 169 | 31 | 18 | 6 | 1817599 | 0.12 |
| 1998 | 82 | 38 | 46 | 8 | 2531043 | 0.08 |
| 1999 | 102 | 37 | 36 | 8 | 1642271 | 0.12 |
| 2000 | 100 | 36 | 36 | 9 | 2570994 | 0.08 |
| 2001 | 109 | 39 | 36 | 9 | 3150674 | 0.14 |
| 2002 | 100 | 44 | 44 | 12 | 3872427 | 0.11 |
| 2003 | 94 | 41 | 44 | 20 | 8739056 | 0.11 |
| 2004 | 94 | 44 | 47 | 8 | 3598436 | 0.10 |
| 2005 | 105 | 40 | 38 | 7 | 3005315 | 0.08 |
| 2006 | 110 | 36 | 33 | 14 | 5518000 | 0.12 |
| 2007 | 103 | 33 | 32 | 8 | 3661314 | 0.14 |
| 2008 | 105 | 40 | 38 | 10 | 6468839 | 0.15 |
| 2009 | 102 | 26 | 26 | 7 | 2564694 | 0.09 |
| 2010 | 101 | 30 | 30 | 4 | 1804538 | 0.10 |
| 2011 | 108 | 27 | 25 | 4 | 1513742 | 0.12 |
| 2012 | 96 | 25 | 26 | 5 | 2034552 | 0.11 |
| 2013 | 96 | 19 | 20 | 4 | 995987 | 0.13 |
| 2014 | 98 | 20 | 20 | 3 | 669931 | 0.13 |

Table 17. Bss.27.4bc7ad-h: Commercial lpue index for French fleet updated for the 2022 assessment.

| Year | Index |  | +-SE |
| :---: | :---: | :---: | :---: |
|  | 2001 | 0.915 | 0.140 |
|  | 2002 | 0.961 | 0.105 |
|  | 2003 | 0.944 | 0.109 |
|  | 2004 | 0.927 | 0.091 |
|  | 2005 | 1.016 | 0.072 |
|  | 2006 | 1.019 | 0.089 |
|  | 2007 | 1.106 | 0.080 |
|  | 2008 | 1.083 | 0.084 |
|  | 2009 | 1.000 | 0.045 |
|  | 2010 | 0.924 | 0.048 |
|  | 2011 | 0.813 | 0.050 |
|  | 2012 | 0.747 | 0.045 |
|  | 2013 | 0.740 | 0.069 |
|  | 2014 | 0.617 | 0.047 |
|  | 2015 | 0.626 | 0.056 |
|  | 2016 | 0.516 | 0.044 |
|  | 2017 | 0.484 | 0.051 |
|  | 2018 | 0.506 | 0.070 |
|  | 2019 | 0.683 | 0.099 |
|  | 2020 | 0.745 | 0.124 |
|  | 2021 | 0.925 | $0.149$ |

Table 18. Bss.27.4bc7ad-h: Key model assumptions and parameters from the WGCSE 2021 update assessment.

| Characteristic | Settings |
| :--- | :--- |
| Starting year | 1985 |
| Ending year | Assessment year-1 (2021) |
| Equilibrium commercial catch for starting | $0.82^{*}$ landings in 1985 by fleet. |
| year | Constant F estimated using 2012 survey results 1985-2014; 2015-pre- |
| Equilibrium recreational catch for starting | sent Frec multiplier on F 2012 survey results |
| year | Asymptotic, length-based |
| Fumber of areas | 1 |
| Fleet 3: UK Midwater trawl selectivity | As Combined French fleet selectivity |


| Characteristic | Settings |
| :---: | :---: |
| CGFS survey timing (yr) | 0.75 |
| French LPUE timing (yr) | -1 |
| Catchabilities (all surveys) | Analytical solution |
| Survey selectivities: Solent autumn: | Double normal, length-based constrained by Min-Max age selectivity, age-based |
| Survey selectivities: CGFS | Double normal, length-based |
| Tunning fleet: French LPUE | Mirrors French fleet |
| Fixed biological characteristics |  |
| Natural mortality | 0.24 |
| Beverton-Holt steepness | 0.999 |
| Recruitment variability ( $\sigma \mathrm{R}$ ) | 0.9 |
| Weight-length coefficient | 0.00001296 |
| Weight-length exponent | 2.969 |
| Maturity inflection (L50\%) | 40.649 cm |
| Maturity slope | -0.33349 |
| Length-at-age Amin | 19.6 cm at $\mathrm{Amin}=2$ |
| Length-at-Amax | 80.26 cm |
| von Bertalanffy k | 0.09699 |
| von Bertalanffy Linf | 84.55 cm |
| von Bertalanffy t0 | -0.730 yr |
| Std. Deviation length-at-age (cm) | SD $=0.1166$ * age + 3.5609 |
| Age error matrix | CV 12\% at-age |
| Other model settings |  |
| First year for main recruitment deviations | 1955 |
| Last year for recruit deviations | 2018 |
| Last year no bias adjustment | 1973.5 |
| First year full bias adjustment | 1981.7 |
| Last year full bias adjustment | 2018.9 |
| First year recent year no bias adjustment | 2019.7 |
| Maximum bias adjustment | 0.915 |

Table 19. Bss.27.4bc7ad-h: Final seabass update assessment: model estimated stock numbers-at-age (thousands of fish).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 855 | 1199 | 21985 | 8048 | 4913 | 1601 | 1458 | 1220 | 1678 | 4133 | 1443 | 913 | 683 | 522 | 382 | 266 | 589 |
| 1986 | 2490 | 672 | 942 | 17188 | 6221 | 3725 | 1185 | 1054 | 866 | 1179 | 2887 | 1005 | 635 | 475 | 363 | 266 | 594 |
| 1987 | 21289 | 1959 | 528 | 735 | 13246 | 4689 | 2730 | 845 | 736 | 597 | 806 | 1967 | 683 | 431 | 322 | 246 | 584 |
| 1988 | 17384 | 16747 | 1537 | 411 | 563 | 9866 | 3373 | 1898 | 572 | 488 | 392 | 525 | 1276 | 443 | 279 | 209 | 538 |
| 1989 | 90994 | 13675 | 13149 | 1199 | 316 | 423 | 7195 | 2391 | 1317 | 391 | 332 | 265 | 354 | 860 | 298 | 188 | 504 |
| 1990 | 7431 | 71578 | 10737 | 10261 | 923 | 237 | 308 | 5094 | 1657 | 900 | 265 | 224 | 178 | 239 | 580 | 201 | 467 |
| 1991 | 15313 | 5846 | 56198 | 8378 | 7891 | 692 | 173 | 218 | 3522 | 1130 | 609 | 179 | 151 | 120 | 160 | 390 | 450 |
| 1992 | 22686 | 12046 | 4588 | 43797 | 6422 | 5879 | 498 | 120 | 148 | 2352 | 747 | 401 | 117 | 99 | 79 | 105 | 552 |
| 1993 | 8674 | 17845 | 9453 | 3574 | 33570 | 4789 | 4236 | 347 | 82 | 99 | 1552 | 490 | 262 | 77 | 65 | 51 | 430 |
| 1994 | 33431 | 6823 | 14006 | 7368 | 2743 | 25082 | 3460 | 2964 | 237 | 55 | 66 | 1027 | 324 | 173 | 51 | 43 | 318 |
| 1995 | 49741 | 26298 | 5358 | 10932 | 5666 | 2054 | 18173 | 2434 | 2043 | 161 | 37 | 44 | 692 | 218 | 116 | 34 | 244 |
| 1996 | 3124 | 39127 | 20645 | 4178 | 8386 | 4223 | 1477 | 12663 | 1658 | 1372 | 107 | 25 | 29 | 458 | 145 | 77 | 185 |
| 1997 | 57526 | 2457 | 30678 | 16037 | 3180 | 6162 | 2968 | 994 | 8237 | 1053 | 859 | 67 | 15 | 18 | 283 | 89 | 162 |
| 1998 | 17317 | 45252 | 1927 | 23854 | 12226 | 2343 | 4352 | 2013 | 654 | 5303 | 669 | 542 | 42 | 10 | 11 | 178 | 158 |
| 1999 | 56771 | 13622 | 35495 | 1499 | 18212 | 9038 | 1662 | 2967 | 1331 | 423 | 3388 | 425 | 343 | 26 | 6 | 7 | 212 |
| 2000 | 24547 | 44658 | 10683 | 27599 | 1143 | 13428 | 6375 | 1122 | 1934 | 847 | 266 | 2113 | 264 | 213 | 16 | 4 | 137 |
| 2001 | 27708 | 19309 | 35025 | 8308 | 21063 | 845 | 9540 | 4356 | 743 | 1254 | 542 | 169 | 1337 | 167 | 134 | 10 | 89 |
| 2002 | 43508 | 21796 | 15145 | 27242 | 6342 | 15576 | 600 | 6515 | 2883 | 481 | 802 | 344 | 107 | 845 | 105 | 85 | 63 |
| 2003 | 44224 | 34225 | 17097 | 11782 | 20797 | 4688 | 11058 | 410 | 4315 | 1871 | 309 | 511 | 219 | 68 | 536 | 67 | 94 |
| 2004 | 34018 | 34788 | 26828 | 13272 | 8951 | 15239 | 3282 | 7401 | 265 | 2715 | 1158 | 189 | 312 | 133 | 41 | 326 | 98 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 22668 | 26760 | 27266 | 20817 | 10073 | 6546 | 10636 | 2187 | 4751 | 165 | 1670 | 706 | 115 | 189 | 81 | 25 | 257 |
| 2006 | 24147 | 17831 | 20958 | 21107 | 15717 | 7297 | 4504 | 6951 | 1370 | 2885 | 98 | 982 | 413 | 67 | 110 | 47 | 164 |
| 2007 | 27697 | 18995 | 13965 | 16221 | 15929 | 11372 | 5009 | 2937 | 4344 | 830 | 1713 | 58 | 573 | 240 | 39 | 64 | 123 |
| 2008 | 14458 | 21787 | 14883 | 10822 | 12274 | 11575 | 7855 | 3295 | 1857 | 2671 | 501 | 1024 | 34 | 340 | 142 | 23 | 111 |
| 2009 | 12729 | 11373 | 17072 | 11534 | 8186 | 8910 | 7985 | 5163 | 2084 | 1143 | 1615 | 300 | 610 | 20 | 202 | 85 | 80 |
| 2010 | 2456 | 10013 | 8913 | 13237 | 8735 | 5956 | 6171 | 5278 | 3288 | 1293 | 697 | 977 | 181 | 366 | 12 | 121 | 99 |
| 2011 | 10188 | 1932 | 7840 | 6894 | 9967 | 6286 | 4051 | 3974 | 3252 | 1963 | 756 | 403 | 561 | 103 | 210 | 7 | 126 |
| 2012 | 4378 | 8014 | 1513 | 6073 | 5206 | 7212 | 4311 | 2639 | 2484 | 1975 | 1171 | 446 | 237 | 329 | 61 | 123 | 78 |
| 2013 | 15034 | 3444 | 6276 | 1170 | 4565 | 3729 | 4867 | 2752 | 1613 | 1472 | 1148 | 673 | 255 | 135 | 188 | 35 | 115 |
| 2014 | 28374 | 11826 | 2695 | 4839 | 874 | 3224 | 2460 | 3012 | 1619 | 915 | 816 | 628 | 366 | 138 | 73 | 102 | 82 |
| 2015 | 6095 | 22320 | 9269 | 2087 | 3636 | 622 | 2155 | 1561 | 1844 | 969 | 541 | 480 | 369 | 215 | 82 | 43 | 109 |
| 2016 | 25790 | 4795 | 17547 | 7254 | 1598 | 2634 | 416 | 1353 | 949 | 1111 | 584 | 327 | 291 | 225 | 132 | 50 | 95 |
| 2017 | 6281 | 20287 | 3771 | 13770 | 5624 | 1199 | 1879 | 284 | 901 | 628 | 735 | 387 | 218 | 194 | 151 | 89 | 98 |
| 2018 | 19116 | 4941 | 15955 | 2959 | 10683 | 4235 | 866 | 1311 | 195 | 615 | 429 | 503 | 265 | 149 | 134 | 104 | 129 |
| 2019 | 12589 | 15037 | 3886 | 12530 | 2306 | 8146 | 3115 | 614 | 912 | 135 | 425 | 297 | 349 | 185 | 104 | 93 | 163 |
| 2020 | 21771 | 9903 | 11826 | 3051 | 9754 | 1756 | 5997 | 2219 | 430 | 634 | 94 | 296 | 207 | 244 | 129 | 73 | 180 |
| 2021 | 21754 | 17125 | 7787 | 9280 | 2373 | 7421 | 1290 | 4252 | 1542 | 297 | 438 | 65 | 205 | 144 | 169 | 90 | 177 |
| 2022 | 21755 | 17112 | 13467 | 6111 | 7222 | 1808 | 5466 | 917 | 2960 | 1067 | 205 | 303 | 45 | 142 | 100 | 118 | 186 |

Table 20. Bss.27.4bc7ad-h: Final seabass update assessment: model estimated fishing mortality-at-age.

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0.000 | 0.001 | 0.006 | 0.017 | 0.037 | 0.061 | 0.085 | 0.102 | 0.113 | 0.119 | 0.122 | 0.123 | 0.124 | 0.124 | 0.124 | 0.124 | 0.123 |
| 1986 | 0.000 | 0.002 | 0.007 | 0.021 | 0.043 | 0.071 | 0.098 | 0.119 | 0.132 | 0.140 | 0.144 | 0.146 | 0.147 | 0.147 | 0.147 | 0.147 | 0.146 |
| 1987 | 0.000 | 0.003 | 0.010 | 0.027 | 0.055 | 0.089 | 0.123 | 0.151 | 0.170 | 0.182 | 0.189 | 0.192 | 0.194 | 0.195 | 0.195 | 0.195 | 0.194 |
| 1988 | 0.000 | 0.002 | 0.008 | 0.022 | 0.046 | 0.076 | 0.104 | 0.125 | 0.139 | 0.147 | 0.152 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.152 |
| 1989 | 0.000 | 0.002 | 0.008 | 0.022 | 0.047 | 0.077 | 0.105 | 0.127 | 0.141 | 0.148 | 0.153 | 0.154 | 0.155 | 0.155 | 0.155 | 0.154 | 0.152 |
| 1990 | 0.000 | 0.002 | 0.008 | 0.023 | 0.047 | 0.078 | 0.107 | 0.129 | 0.143 | 0.151 | 0.155 | 0.157 | 0.157 | 0.157 | 0.157 | 0.156 | 0.154 |
| 1991 | 0.000 | 0.002 | 0.009 | 0.026 | 0.054 | 0.090 | 0.123 | 0.148 | 0.164 | 0.173 | 0.178 | 0.180 | 0.181 | 0.181 | 0.180 | 0.180 | 0.177 |
| 1992 | 0.000 | 0.002 | 0.010 | 0.026 | 0.053 | 0.088 | 0.121 | 0.147 | 0.165 | 0.175 | 0.181 | 0.184 | 0.185 | 0.186 | 0.186 | 0.185 | 0.184 |
| 1993 | 0.000 | 0.002 | 0.009 | 0.025 | 0.051 | 0.085 | 0.117 | 0.142 | 0.158 | 0.168 | 0.173 | 0.176 | 0.177 | 0.177 | 0.177 | 0.177 | 0.175 |
| 1994 | 0.000 | 0.002 | 0.008 | 0.023 | 0.049 | 0.082 | 0.111 | 0.132 | 0.145 | 0.151 | 0.154 | 0.155 | 0.155 | 0.155 | 0.154 | 0.153 | 0.150 |
| 1995 | 0.000 | 0.002 | 0.009 | 0.025 | 0.054 | 0.090 | 0.121 | 0.144 | 0.158 | 0.166 | 0.170 | 0.171 | 0.171 | 0.171 | 0.170 | 0.169 | 0.166 |
| 1996 | 0.000 | 0.003 | 0.013 | 0.033 | 0.068 | 0.113 | 0.156 | 0.190 | 0.214 | 0.228 | 0.237 | 0.241 | 0.242 | 0.243 | 0.243 | 0.242 | 0.240 |
| 1997 | 0.000 | 0.003 | 0.012 | 0.031 | 0.065 | 0.108 | 0.148 | 0.179 | 0.200 | 0.213 | 0.220 | 0.223 | 0.224 | 0.224 | 0.224 | 0.223 | 0.220 |
| 1998 | 0.000 | 0.003 | 0.011 | 0.030 | 0.062 | 0.103 | 0.143 | 0.174 | 0.195 | 0.208 | 0.215 | 0.218 | 0.220 | 0.220 | 0.220 | 0.219 | 0.217 |
| 1999 | 0.000 | 0.003 | 0.012 | 0.031 | 0.065 | 0.109 | 0.154 | 0.188 | 0.211 | 0.225 | 0.232 | 0.236 | 0.237 | 0.237 | 0.237 | 0.237 | 0.234 |
| 2000 | 0.000 | 0.003 | 0.011 | 0.030 | 0.062 | 0.102 | 0.141 | 0.172 | 0.193 | 0.206 | 0.214 | 0.218 | 0.219 | 0.220 | 0.219 | 0.219 | 0.217 |
| 2001 | 0.000 | 0.003 | 0.011 | 0.030 | 0.062 | 0.102 | 0.141 | 0.173 | 0.194 | 0.207 | 0.214 | 0.218 | 0.219 | 0.220 | 0.220 | 0.219 | 0.217 |
| 2002 | 0.000 | 0.003 | 0.011 | 0.030 | 0.062 | 0.103 | 0.142 | 0.172 | 0.193 | 0.205 | 0.211 | 0.214 | 0.216 | 0.216 | 0.215 | 0.215 | 0.212 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 0.000 | 0.003 | 0.013 | 0.035 | 0.071 | 0.116 | 0.162 | 0.198 | 0.223 | 0.239 | 0.248 | 0.253 | 0.255 | 0.255 | 0.255 | 0.255 | 0.252 |
| 2004 | 0.000 | 0.004 | 0.014 | 0.036 | 0.073 | 0.120 | 0.166 | 0.203 | 0.229 | 0.246 | 0.255 | 0.260 | 0.262 | 0.262 | 0.262 | 0.262 | 0.259 |
| 2005 | 0.000 | 0.004 | 0.016 | 0.041 | 0.082 | 0.134 | 0.185 | 0.228 | 0.259 | 0.279 | 0.291 | 0.297 | 0.300 | 0.301 | 0.301 | 0.301 | 0.298 |
| 2006 | 0.000 | 0.004 | 0.016 | 0.041 | 0.084 | 0.136 | 0.188 | 0.230 | 0.261 | 0.281 | 0.293 | 0.299 | 0.302 | 0.303 | 0.303 | 0.303 | 0.300 |
| 2007 | 0.000 | 0.004 | 0.015 | 0.039 | 0.079 | 0.130 | 0.179 | 0.219 | 0.247 | 0.265 | 0.275 | 0.280 | 0.282 | 0.283 | 0.283 | 0.282 | 0.279 |
| 2008 | 0.000 | 0.004 | 0.015 | 0.039 | 0.080 | 0.131 | 0.180 | 0.218 | 0.245 | 0.263 | 0.273 | 0.277 | 0.279 | 0.280 | 0.279 | 0.278 | 0.275 |
| 2009 | 0.000 | 0.004 | 0.014 | 0.038 | 0.078 | 0.127 | 0.174 | 0.211 | 0.237 | 0.254 | 0.263 | 0.268 | 0.270 | 0.270 | 0.270 | 0.269 | 0.265 |
| 2010 | 0.000 | 0.005 | 0.017 | 0.044 | 0.089 | 0.145 | 0.200 | 0.244 | 0.276 | 0.297 | 0.309 | 0.315 | 0.317 | 0.318 | 0.318 | 0.317 | 0.314 |
| 2011 | 0.000 | 0.004 | 0.015 | 0.041 | 0.084 | 0.137 | 0.189 | 0.230 | 0.258 | 0.277 | 0.287 | 0.292 | 0.294 | 0.294 | 0.293 | 0.293 | 0.288 |
| 2012 | 0.000 | 0.004 | 0.017 | 0.045 | 0.094 | 0.153 | 0.209 | 0.253 | 0.283 | 0.302 | 0.313 | 0.318 | 0.320 | 0.320 | 0.319 | 0.318 | 0.312 |
| 2013 | 0.000 | 0.005 | 0.020 | 0.053 | 0.108 | 0.176 | 0.240 | 0.291 | 0.327 | 0.351 | 0.364 | 0.370 | 0.372 | 0.372 | 0.372 | 0.370 | 0.364 |
| 2014 | 0.000 | 0.004 | 0.015 | 0.046 | 0.100 | 0.163 | 0.215 | 0.251 | 0.273 | 0.285 | 0.290 | 0.291 | 0.290 | 0.287 | 0.284 | 0.281 | 0.269 |
| 2015 | 0.000 | 0.001 | 0.005 | 0.027 | 0.083 | 0.161 | 0.226 | 0.258 | 0.267 | 0.267 | 0.264 | 0.260 | 0.256 | 0.251 | 0.247 | 0.243 | 0.224 |
| 2016 | 0.000 | 0.000 | 0.002 | 0.014 | 0.047 | 0.098 | 0.144 | 0.167 | 0.173 | 0.172 | 0.170 | 0.167 | 0.163 | 0.160 | 0.156 | 0.153 | 0.139 |
| 2017 | 0.000 | 0.000 | 0.002 | 0.014 | 0.044 | 0.086 | 0.120 | 0.136 | 0.141 | 0.141 | 0.140 | 0.138 | 0.136 | 0.134 | 0.132 | 0.130 | 0.121 |
| 2018 | 0.000 | 0.000 | 0.002 | 0.010 | 0.031 | 0.067 | 0.104 | 0.123 | 0.129 | 0.129 | 0.128 | 0.126 | 0.124 | 0.122 | 0.120 | 0.118 | 0.109 |
| 2019 | 0.000 | 0.000 | 0.002 | 0.010 | 0.032 | 0.066 | 0.099 | 0.117 | 0.123 | 0.123 | 0.122 | 0.121 | 0.119 | 0.118 | 0.116 | 0.115 | 0.108 |
| 2020 | 0.000 | 0.000 | 0.002 | 0.011 | 0.033 | 0.068 | 0.104 | 0.124 | 0.130 | 0.131 | 0.130 | 0.128 | 0.127 | 0.125 | 0.123 | 0.122 | 0.115 |
| 2021 | 0.000 | 0.000 | 0.002 | 0.011 | 0.032 | 0.066 | 0.102 | 0.122 | 0.129 | 0.130 | 0.129 | 0.127 | 0.125 | 0.124 | 0.122 | 0.120 | 0.113 |

Table 21. Bss.27.4bc7ad-h: Final seabass update assessment: stock summary table.

| Year | Low | Recruitment (Age 0, thousands) | High | Low | SSB <br> (Tonnes) | High | Low | $\begin{aligned} & \text { F(4- } \\ & \text { 15) } \end{aligned}$ | High | $\mathrm{F}_{\text {commer- }}$ cial | $F_{\text {recrea- }}$ <br> tional | Commercial landings | Commercial discards* | Recreational removals |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 74 | 855 | 1635 | 17572 | 23629 | 29686 | 0.076 | 0.105 | 0.133 | 0.037 | 0.068 | 994 |  | 1820 |
| 1986 | 442 | 2490 | 4538 | 15315 | 20735 | 26155 | 0.09 | 0.123 | 0.157 | 0.056 | 0.068 | 1318 |  | 1638.5 |
| 1987 | 15045 | 21289 | 27534 | 13658 | 18496 | 23333 | 0.117 | 0.161 | 0.2 | 0.093 | 0.068 | 1979 |  | 1493 |
| 1988 | 9976 | 17384 | 24793 | 12431 | 16800 | 21168 | 0.095 | 0.13 | 0.165 | 0.062 | 0.068 | 1239 |  | 1384 |
| 1989 | 76439 | 90994 | 105549 | 12061 | 16146 | 20231 | 0.096 | 0.131 | 0.166 | 0.063 | 0.067 | 1161 |  | 1279 |
| 1990 | 2289 | 7431 | 12574 | 10912 | 14789 | 18665 | 0.095 | 0.133 | 0.17 | 0.066 | 0.068 | 1064 |  | 1150 |
| 1991 | 9230 | 15313 | 21397 | 9323 | 12945 | 16566 | 0.108 | 0.153 | 0.197 | 0.085 | 0.068 | 1226 |  | 1056 |
| 1992 | 15613 | 22686 | 29758 | 8027 | 11350 | 14673 | 0.111 | 0.155 | 0.199 | 0.087 | 0.068 | 1186 |  | 1080 |
| 1993 | 4094 | 8674 | 13254 | 8663 | 11735 | 14807 | 0.111 | 0.148 | 0.185 | 0.08 | 0.068 | 1256 |  | 1258 |
| 1994 | 23880 | 33431 | 42982 | 11432 | 14377 | 17322 | 0.105 | 0.133 | 0.161 | 0.066 | 0.067 | 1370 |  | 1509 |
| 1995 | 38898 | 49741 | 60583 | 15017 | 18087 | 21156 | 0.118 | 0.146 | 0.175 | 0.079 | 0.067 | 1835 |  | 1689 |
| 1996 | 455 | 3124 | 5793 | 16747 | 20021 | 23296 | 0.161 | 0.2 | 0.24 | 0.133 | 0.068 | 3022 |  | 1696 |
| 1997 | 45330 | 57526 | 69722 | 15828 | 19147 | 22467 | 0.15 | 0.188 | 0.23 | 0.120 | 0.068 | 2620 |  | 1605 |
| 1998 | 8068 | 17317 | 26565 | 14607 | 17868 | 21128 | 0.145 | 0.183 | 0.22 | 0.115 | 0.068 | 2390 |  | 1548 |
| 1999 | 43395 | 56771 | 70148 | 14047 | 17205 | 20362 | 0.156 | 0.197 | 0.24 | 0.129 | 0.068 | 2670 |  | 1548 |
| 2000 | 15298 | 24547 | 33797 | 14165 | 17255 | 20346 | 0.144 | 0.182 | 0.22 | 0.114 | 0.068 | 2407 |  | 1601 |
| 2001 | 15645 | 27708 | 39770 | 15049 | 18200 | 21351 | 0.145 | 0.182 | 0.22 | 0.114 | 0.068 | 2500 |  | 1685 |
| 2002 | 28707 | 43508 | 58309 | 15800 | 19024 | 22247 | 0.143 | 0.18 | 0.22 | 0.112 | 0.068 | 2622 | 17 | 1782 |
| 2003 | 31439 | 44224 | 57010 | 16902 | 20213 | 23524 | 0.168 | 0.21 | 0.25 | 0.143 | 0.068 | 3459 | 16 | 1863 |
| 2004 | 23245 | 34018 | 44791 | 17568 | 20953 | 24337 | 0.172 | 0.22 | 0.26 | 0.149 | 0.068 | 3731 | 59 | 1908 |
| 2005 | 14857 | 22668 | 30479 | 18014 | 21470 | 24925 | 0.194 | 0.25 | 0.3 | 0.178 | 0.068 | 4430 | 96 | 1905 |
| 2006 | 16681 | 24147 | 31614 | 17485 | 20993 | 24501 | 0.195 | 0.25 | 0.3 | 0.180 | 0.068 | 4377 | 53 | 1874 |
| 2007 | 18925 | 27697 | 36469 | 17050 | 20524 | 23998 | 0.185 | 0.23 | 0.28 | 0.165 | 0.068 | 4064 | 50 | 1871 |


| Year | Low | Recruitment (Age 0, thousands) | High | Low | SSB <br> (Tonnes) | High | Low | $\begin{aligned} & \text { F(4- } \\ & 15) \end{aligned}$ | High | $F_{\text {commer- }}$ cial | $F_{\text {recrea- }}$ <br> tional | Commercial landings | Commercial discards* | Recreational removals |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 7871 | 14458 | 21046 | 17419 | 20797 | 24175 | 0.184 | 0.23 | 0.28 | 0.164 | 0.068 | 4107 | 8 | 1878 |
| 2009 | 8033 | 12729 | 17424 | 17793 | 21070 | 24347 | 0.18 | 0.22 | 0.27 | 0.156 | 0.068 | 3889 | 151.2 | 1852 |
| 2010 | 339 | 2456 | 4572 | 17643 | 20805 | 23967 | 0.21 | 0.26 | 0.31 | 0.194 | 0.068 | 4562 | 147.9 | 1751 |
| 2011 | 6762 | 10188 | 13613 | 16213 | 19192 | 22171 | 0.196 | 0.24 | 0.29 | 0.176 | 0.068 | 3858 | 22 | 1605 |
| 2012 | 2163 | 4378 | 6593 | 14932 | 17702 | 20471 | 0.21 | 0.27 | 0.32 | 0.199 | 0.068 | 3987 | 156.6 | 1440 |
| 2013 | 9443 | 15034 | 20624 | 13144 | 15749 | 18355 | 0.24 | 0.31 | 0.38 | 0.24 | 0.068 | 4137 | 53.4 | 1222 |
| 2014 | 18802 | 28374 | 37947 | 10524 | 13040 | 15557 | 0.191 | 0.25 | 0.31 | 0.184 | 0.067 | 2682 | 24.7 | 1008 |
| 2015 | 2030 | 6095 | 10161 | 8483 | 10967 | 13451 | 0.172 | 0.23 | 0.29 | 0.177 | 0.055 | 2066 | 39.5 | 689.5 |
| 2016 | 13317 | 25790 | 38263 | 6603 | 9058 | 11514 | 0.104 | 0.147 | 0.191 | 0.128 | 0.019 | 1295 | 198.6 | 209 |
| 2017 | 1860 | 6281 | 10702 | 5810 | 8254 | 10699 | 0.085 | 0.123 | 0.161 | 0.104 | 0.0189 | 984 | 271.102 | 206 |
| 2018 | 5906 | 19116 | 32326 | 5706 | 8231 | 10756 | 0.074 | 0.11 | 0.145 | 0.097 | 0.0131 | 948 | 482.4 | 150 |
| 2019 | 2009 | 12589 | 23169 | 6251 | 9033 | 11816 | 0.072 | 0.106 | 0.14 | 0.085 | 0.021 | 972 | 463.9 | 274 |
| 2020 |  | 10105** |  | 7130 | 10369 | 13609 | 0.075 | 0.112 | 0.15 | 0.081 | 0.032 | 1042 | 325 | 453 |
| 2021 |  | 10105** |  | 7754 | 11491 | 15227 | 0.072 | 0.111 | 0.149 | 0.079 | 0.032 | 1126 | 411.8 | 489 |
| 2022 |  | 10105** |  | 8128 | 12384 | 16640 |  |  |  |  |  |  |  |  |

* Incomplete for some fleets 2002-2008.
**Geometric mean recruitment 2010-2019.

Table 22. Bss.27.4bc7ad-h: Inputs for short-term forecast. Fishing mortality is the estimates for 2022. Numbers-at-ages 0-2 in $\mathbf{2 0 2 2}$ are adjusted by replacing Stock Synthesis values for 0-group in 2020-2021 (years with no recruit deviations estimated) with the long-term GM, adjusted for natural mortality.

| age | Stock <br> num- <br> bers <br> 2022 | weight in stock | Pro- <br> por- <br> tion <br> ma- <br> ture <br> (fe- <br> male) | H.Cons retained mean F | H.Cons Discarded mean F | H.Cons retained mean weights | H.Cons discarded mean weights | H.Cons <br> pro- <br> por- <br> tion <br> re- <br> tained | Rec-reational F | Recre- <br> a- <br> tional <br> re- <br> mov- <br> als <br> mean <br> weight | M |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 10105 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.020 | 0.24 |
| 1 | 7949 | 0.024 | 0.000 | 0.000 | 0.000 | 0.117 | 0.117 | 0.188 | 0.000 | 0.079 | 0.24 |
| 2 | 6251 | 0.096 | 0.000 | 0.000 | 0.001 | 0.236 | 0.237 | 0.190 | 0.002 | 0.191 | 0.24 |
| 3 | 6111 | 0.210 | 0.000 | 0.001 | 0.005 | 0.385 | 0.385 | 0.209 | 0.005 | 0.342 | 0.24 |
| 4 | 7222 | 0.369 | 0.093 | 0.007 | 0.014 | 0.588 | 0.562 | 0.338 | 0.012 | 0.531 | 0.24 |
| 5 | 1808 | 0.571 | 0.297 | 0.030 | 0.019 | 0.812 | 0.753 | 0.634 | 0.020 | 0.751 | 0.24 |
| 6 | 5466 | 0.808 | 0.578 | 0.066 | 0.012 | 1.014 | 0.967 | 0.853 | 0.027 | 0.993 | 0.24 |
| 7 | 917 | 1.073 | 0.797 | 0.088 | 0.005 | 1.243 | 1.221 | 0.949 | 0.031 | 1.252 | 0.24 |
| 8 | 2960 | 1.358 | 0.914 | 0.095 | 0.002 | 1.512 | 1.505 | 0.984 | 0.034 | 1.528 | 0.24 |
| 9 | 1067 | 1.658 | 0.965 | 0.096 | 0.000 | 1.804 | 1.806 | 0.995 | 0.035 | 1.821 | 0.24 |
| 10 | 205 | 1.965 | 0.985 | 0.095 | 0.000 | 2.106 | 2.113 | 0.998 | 0.035 | 2.124 | 0.24 |
| 11 | 303 | 2.276 | 0.993 | 0.093 | 0.000 | 2.410 | 2.420 | 0.999 | 0.035 | 2.432 | 0.24 |
| 12 | 45 | 2.584 | 0.997 | 0.091 | 0.000 | 2.709 | 2.722 | 0.999 | 0.035 | 2.737 | 0.24 |
| 13 | 142 | 2.888 | 0.998 | 0.090 | 0.000 | 3.002 | 3.018 | 1.000 | 0.035 | 3.037 | 0.24 |
| 14 | 100 | 3.183 | 0.999 | 0.088 | 0.000 | 3.286 | 3.305 | 1.000 | 0.036 | 3.327 | 0.24 |
| 15 | 118 | 3.468 | 0.999 | 0.086 | 0.000 | 3.559 | 3.579 | 1.000 | 0.036 | 3.607 | 0.24 |
| 16 | 186 | 4.184 | 1.000 | 0.085 | 0.000 | 4.216 | 4.063 | 1.000 | 0.036 | 3.874 | 0.24 |

Age 0,1,2 over-written as follows:
2022 yc 2022 age 0 replaced by 20010-2019 LTGM (10104);
2021 yc 2022 age 1 from SS3 survivor estimate at-age 1, 2022 * LTGM / SS3 estimate of age 0 in 2021;
2020 yc 2022 age 2 from SS3 survivor estimate at-age 2, 2022 * LTGM / SS3 estimate of age 0 in 2020.

Table 23. Bss.27.4bc7ad-h: Management options table. F-Multipliers for 2023 are applied to both the commercial and recreational fishery. Note that the combined total commercial and recreational forecasted catch could be allocated in different ways (see advice sheet for footnotes).

| Basis | Total removals $\bullet(2023)$ | $\left\{\begin{array}{c} \text { Projecte } \\ \text { d } \\ \text { landings } \\ (2023) \end{array}\right.$ | $\left\|\begin{array}{c} \text { Projecte } \\ d \\ \text { discards } \\ (2023) \end{array}\right\|$ | Recreation al removals (2023) | $\underset{(2023)}{E_{1} \operatorname{sos}}$ | $\left\|\begin{array}{c} \text { Fescivciof } \\ \text { Lardne } \\ (2023) \end{array}\right\|$ | Fevivert doxat (2023) | F <br> recreation <br> al <br> removals <br> (2023) | $\frac{\text { SSB }}{(2024)}$ | \% SSB change ${ }^{* *}$ | \% advice <br> change *** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |  |  |  |
| MSY approach: <br> $\mathrm{F}_{\mathrm{MSY}} \times \mathrm{SSB}_{\text {2023 }} / \mathrm{MSY}^{\text {gunss }}$ | 2542 | 1653 | 194 | 695 | 0.153 | 0.105 | 0.0059 | 0.042 | 12631 | -2.1 | 14.7 |
|  |  |  |  |  |  |  |  |  |  |  |  |
| EU MAPA: | 2542 | 1653 | 194 | 695 | 0.153 | 0.105 | 0.0059 | 0.042 | 12631 | -2.1 | 14.7 |
| EU MAPA: <br> $\mathrm{F}_{\text {MSY и црре }} \times \mathrm{SSB}_{2023} / \mathrm{MSY}_{\mathrm{B}_{\text {ti }}}$ | 2542 | 1653 | 194 | 695 | 0.153 | 0.105 | 0.0059 | 0.042 | 12631 | -2.1 | 14.7 |
| EU MAPA: <br> $\mathrm{F}_{\text {MSY lewt }} \times$ SSB $_{2023} /$ MSY B $_{\text {Buth }}$选 | 2133 | 1388 | 162 | 583 | 0.127 | 0.087 | 0.0049 | 0.035 | 12964 | 0.50 | -3.7 |
| $\mathrm{F}=\mathrm{F}_{\text {MSY }}$ | 2821 | 1834 | 216 | 771 | 0.171 | 0.117 | 0.0066 | 0.047 | 12405 | -3.8 | 27 |
| $\mathrm{F}=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14716 | 14.1 | -100 |
| $\mathrm{F}_{06}$ | 3294 | 2140 | 253 | 901 | 0.20 | 0.139 | 0.0079 | 0.056 | 12021 | -6.8 | 49 |
| $\mathrm{F}_{\text {dus }}$ | 4096 | 2612 | 313 | 1101 | 0.25 | 0.174 | 0.0098 | 0.070 | 11430 | -11.4 | 85 |
| $\mathrm{SSB}_{2024}=\mathrm{B}_{0 \times 1}$ | 5422 | 3509 | 429 | 1484 | 0.36 | 0.24 | 0.0139 | 0.099 | 10313 | -20 | 145 |
| $\mathrm{SSB}_{2024}=\mathrm{B}_{3}$ | 336 | 219 | 25 | 92 | 0.0190 | 0.0130 | 0.000 | 0.005 | 14439 | 11.9 | -85 |
|  | 336 | 219 | 25 | 92 | 0.0190 | 0.0130 | 0.000 | 0.005 | 14439 | 11.9 | -85 |
| $\mathrm{F}=\mathrm{F}_{2023}$ | 1906 | 1241 | 144 | 521 | 0.113 | 0.077 | 0.0044 | 0.031 | 13149 | 1.94 | -14.0 |
| $\mathrm{SSB}_{2024}=$ SSB $_{2003}$ | 2213 | 1440 | 168 | 605 | 0.132 | 0.090 | 0.0051 | 0.036 | 12899 | 0.00 | -0.135 |



Figure 1. Bss.27.4bc7ad-h: Trends in official sea bass landings by country.


Figure 2. Bss.27.4bc7ad-h: Trends in ICES estimates of seabass landings by gear (France -top- and UK -bottom).


Figure 3. Bss.27.4bc7ad-h: Length composition for the combined French fleet from 2000 onwards.


Figure 4. Bss.27.4bc7ad-h: Length composition of UK bottom trawls and nets fleet landings from 1985 onwards


Figure 5. Bss.27.4bc7ad-h: Length composition of UK Lines fleet landings from 1985 onwards.


Figure 6. Bss.27.4bc7ad-h: Available length composition of UK Midwater pair trawl fleet landings.


Figure 7. Bss.27.4bc7ad-h: Location of Cefas Solent and Thames juvenile seabass surveys.


Figure 8. Bss.27.4bc7ad-h: Left: stations fished during the Channel Groundfish Survey carried out annually by France. Right: distribution of total catches of seabass over the survey series.


Figure 9. Bss.27.4bc7ad-h: Comparison of French commercial LPUE index for European seabass in ICES divisions 4bc and 7a,d-h between last year's assessment and the updated 2022 LPUE.


Figure 10. Bss.27.4bc7ad-h: Top: Datasets used in the updated assessment. Bottom: Landings series for the six fleets.



Ending year selectivity for Other


Ending year selectivity for AutBass


Ending year selectivity for FR_LPUE


Ending year selectivity for RecFish


Ending year selectivity for CGFS1


Ending year selectivity for AutBass

Figure 11. Bss.27.4bc7ad-h: Final seabass update assessment: Fitted length-based and age-based selectivity curves.


Surface plot of Time-varying selectivity for UKOTB_Nets Surface plot of Time-varying retention for UKOTB_Nets


Surface plot of Time-varying selectivity for Lines


Surface plot of Time-varying selectivity for French Surface plot of Time-varying retention for French


Surface plot of Time-varying selectivity for FR_LPUE Surface plot of Time-varying selectivity for Other

Figure 12. Bss.27.4bc7ad-h: Final seabass update assessment: Fitted time-series of length-based and age-based selectivity and retention curves for fleets with blocks.





Figure 13. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of UK trawl and net fishery-length composition data for the retained (top 3) and discarded (bottom 2) catch components.


Figure 14. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of UK lines length-composition data for the retained catch components.


Figure 15. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of UK midwater trawl fishery lengthcomposition data for the retained catch components.


Figure 16. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of French fishery length-composition data for the retained (top row) and discarded (bottom row) catch components.



Figure 17. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of recreational length-compositions data.


Figure 18. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of Channel groundfish survey lengthcompositions.


Figure 19. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of the commercial fisheries and Channel groundfish survey length compositions, aggregated across time for the retained and discarded catch components.


Figure 20. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of age composition data for the combined UK otter trawl and nets fleets.



Figure 21. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of age composition data for the combined UK lines fleet.


Figure 22. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of age composition data for the UK midwater trawl fleet.


Figure 23. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of age composition data for the combined French fleets.



Figure 24. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of age composition data for the Solent Autumn bass survey.


Year

Figure 25. Bss.27.4bc7ad-h: Final seabass update assessment: Fit and residuals of UK fleets age compositions, aggregated across time.


Figure 26. Bss.27.4bc7ad-h: Final seabass update assessment: Fit to Solent Autumn bass survey total abundance index, accounting for age and length-based selectivity.


Figure 27. Bss.27.4bc7ad-h: Final seabass update assessment: Fit to Channel groundfish survey total abundance index, accounting for length-based selectivity.


Figure 28. Bss.27.4bc7ad-h: Final seabass update assessment: Fit to the French landings per unit of effort commercial index, accounting for length-based selectivity.


Figure 29. Bss.27.4bc7ad-h: Final seabass update assessment: Top: time-series of log-recruit deviations (deviations for 1965-1984 precede the period of input catch data). Below: stock-recruit scatter (model is fitted assuming BevertonHolt stock-recruit model and steepness $=0.999$ ).


Figure 30. Bss.27.4bc7ad-h: Retrospective analysis of stock trends from final update assessment, based on Stock Synthesis run final year set to 2021 and peeling back five years (for the final run, terminal F is for 2020 and SSB and total biomass terminate in 2021).


Figure 31. Bss.27.4bc7ad-h: Stock trends from final update assessment, based on Stock Synthesis run final year set at 2021 to give 2022 numbers and biomass and 2021 F. Recruitment in 2020-2022 is the geometric mean 2010-2019. Recruitment, $F$ and SSB are shown with $95 \%$ confidence intervals.


Figure 32. Bss.27.4bc7ad-h: Comparison between LPUE and recreational catch time-series from this year's final update assessment and the 2021 WGCSE assessment.


Figure 33. Bss.27.4bc7ad-h: Comparison between stock trends from this year's final update assessment and last year WGCSE assessment.

## 30 Sole (Solea solea) in divisions 7.b and 7.c (West of Ireland)

## Type of assessment in 2022

No assessment was performed.

### 30.1 General

### 30.1.1 Stock identity

Sole in 7.b are mainly caught by Irish vessels on sandy grounds in coastal areas. Sole catches in 7.c are negligible. In 7.b there are two distinct areas where sole are caught: an area around Galway Bay and an area in the north of 7.b, which extends into 6.a (the Stags and Broadhaven Ground). The landings and LPUE of sole in 7.bc appear to have been more or less stable since the start of the logbooks' time-series in 1995. It is not known how much exchange there is between sole on the Aran Grounds and those on the Stags and Broadhaven Ground.

### 30.1.2 Data

The time-series of official landings is presented in Table 30.1 and Figure 30.1.
The time-series of otter-trawl landings effort and LPUE since 1995 are shown in Figure 30.2. Landings and effort have gradually declined since the late 1990s and early 2000s, giving rise to relatively stable LPUE over the time-series, with fluctuations.

### 30.1.3 Historical stock development

No analytical assessment was performed.

Table 30.1. Landings of Sole in 7.bc as officially reported to ICES.

| Year | BEL | FRA | UK | IRL | OTH | TOT | Year | BEL | FRA | UK | IRL | OTH | TOT | Unalloc | WG est |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1908 | 0 | 0 | 1 | 37 | 0 | 38 | 1965 | 0 | 95 | 5 | 24 | 0 | 124 |  |  |
| 1909 | 0 | 0 | 0 | 32 | 0 | 32 | 1966 | 0 | 0 | 1 | 11 | 0 | 12 |  |  |
| 1910 | 0 | 0 | 0 | 28 | 0 | 28 | 1967 | 0 | 78 | 0 | 11 | 0 | 89 |  |  |
| 1911 | 0 | 0 | 1 | 22 | 0 | 23 | 1968 | 0 | 121 | 0 | 8 | 0 | 129 |  |  |
| 1912 | 0 | 0 | 1 | 22 | 0 | 23 | 1969 | 0 | 86 | 1 | 9 | 0 | 96 |  |  |
| 1913 | 0 | 0 | 1 | 25 | 0 | 26 | 1970 | 0 | 3 | 0 | 8 | 0 | 11 |  |  |
| 1914 | 0 | 0 | 1 | 43 | 0 | 44 | 1971 | 0 | 0 | 2 | 5 | 0 | 7 |  |  |
| 1915 | 0 | 0 | 1 | 12 | 0 | 13 | 1972 | 0 | 4 | 0 | 13 | 0 | 17 |  |  |
| 1916 | 0 | 0 | 0 | 14 | 0 | 14 | 1973 | 0 | 0 | 0 | 12 | 0 | 12 |  |  |
| 1917 | 0 | 0 | 0 | 6 | 0 | 6 | 1974 | 0 | 25 | 0 | 12 | 0 | 37 |  |  |
| 1918 | 0 | 0 | 0 | 7 | 0 | 7 | 1975 | 0 | 7 | 0 | 19 | 0 | 26 |  |  |
| 1919 | 0 | 0 | 0 | 6 | 0 | 6 | 1976 | 0 | 6 | 0 | 44 | 0 | 50 |  |  |
| 1920 | 0 | 0 | 9 | 5 | 0 | 14 | 1977 | 0 | 3 | 0 | 14 | 0 | 17 |  |  |
| 1921 | 0 | 0 | 10 | 9 | 0 | 19 | 1978 | 0 | 3 | 0 | 16 | 0 | 19 |  |  |
| 1922 | 0 | 0 | 4 | 9 | 0 | 13 | 1979 | 0 | 6 | 0 | 13 | 0 | 19 |  |  |
| 1923 | 0 | 0 | 2 | 10 | 0 | 12 | 1980 | 0 | 9 | 0 | 24 | 0 | 33 |  |  |
| 1924 | 0 | 0 | 15 | 64 | 0 | 79 | 1981 | 0 | 6 | 0 | 47 | 0 | 53 |  |  |
| 1925 | 0 | 0 | 11 | 18 | 0 | 29 | 1982 | 0 | 5 | 1 | 55 | 0 | 61 |  |  |
| 1926 | 0 | 7 | 10 | 18 | 0 | 35 | 1983 | 0 | 9 | 0 | 40 | 0 | 49 |  |  |
| 1927 | 0 | 47 | 11 | 19 | 0 | 77 | 1984 | 0 | 3 | 0 | 17 | 0 | 20 |  |  |
| 1928 | 0 | 49 | 8 | 16 | 0 | 73 | 1985 | 0 | 6 | 0 | 44 | 0 | 50 |  |  |
| 1929 | 0 | 74 | 11 | 18 | 0 | 103 | 1986 | 0 | 8 | 0 | 29 | 0 | 37 |  |  |
| 1930 | 0 | 52 | 5 | 22 | 0 | 79 | 1987 | 0 | 2 | 0 | 39 | 0 | 41 |  |  |
| 1931 | 0 | 82 | 9 | 29 | 0 | 120 | 1988 | 0 | 2 | 1 | 34 | 0 | 37 |  |  |
| 1932 | 0 | 122 | 10 | 27 | 0 | 159 | 1989 | 0 | 0 | 0 | 38 | 0 | 38 |  |  |
| 1933 | 0 | 411 | 10 | 10 | 0 | 431 | 1990 | 0 | 0 | 0 | 41 | 0 | 41 |  |  |
| 1934 | 0 | 217 | 10 | 13 | 0 | 240 | 1991 | 0 | 5 | 0 | 46 | 0 | 51 |  |  |
| 1935 | 0 | 40 | 7 | 11 | 0 | 58 | 1992 | 0 | 2 | 0 | 43 | 0 | 45 |  |  |
| 1936 | 0 | 43 | 20 | 9 | 0 | 72 | 1993 | 0 | 1 | 0 | 59 | 0 | 60 | 0 | 60 |
| 1937 | 0 | 32 | 25 | 14 | 0 | 71 | 1994 | 0 | 1 | 0 | 60 | 0 | 61 | 9 | 70 |
| 1938 | 0 | 44 | 21 | 7 | 0 | 72 | 1995 | 0 | 2 | 0 | 59 | 0 | 61 | -2 | 59 |
| 1939 | 0 | 0 | 0 | 13 | 0 | 13 | 1996 | 0 | 2 | 0 | 52 | 0 | 54 | 3 | 57 |
| 1940 | 0 | 0 | 0 | 19 | 0 | 19 | 1997 | 0 | 3 | 1 | 51 | 0 | 55 | 0 | 55 |
| 1941 | 0 | 0 | 0 | 14 | 0 | 14 | 1998 | 0 | 0 | 0 | 49 | 0 | 49 | 17 | 66 |
| 1942 | 0 | 0 | 0 | 8 | 0 | 8 | 1999 | 0 | 0 | 0 | 68 | 0 | 68 | 4 | 72 |
| 1943 | 0 | 0 | 0 | 11 | 0 | 11 | 2000 | 0 | 12 | 0 | 65 | 0 | 77 | -9 | 68 |
| 1944 | 0 | 0 | 0 | 16 | 0 | 16 | 2001 | 0 | 7 | 0 | 53 | 0 | 60 | 0 | 60 |
| 1945 | 0 | 0 | 0 | 20 | 0 | 20 | 2002 | 0 | 14 | 0 | 50 | 0 | 64 | -3 | 61 |
| 1946 | 0 | 0 | 12 | 10 | 0 | 22 | 2003 | 0 | 19 | 0 | 50 | 0 | 69 | -5 | 64 |


| Year | BEL | FRA | UK | IRL | OTH | TOT | Year | BEL | FRA | UK | IRL | OTH | TOT | Unalloc | WG est |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1947 | 15 | 0 | 6 | 8 | 0 | 29 | 2004 | 0 | 18 | 0 | 49 | 0 | 67 | 2 | 69 |
| 1948 | 0 | 0 | 11 | 14 | 0 | 25 | 2005 | 0 | 7 | 0 | 38 | 0 | 45 | -1 | 44 |
| 1949 | 0 | 41 | 12 | 12 | 0 | 65 | 2006 | 0 | 12 | 0 | 31 | 0 | 43 | 1 | 43 |
| 1950 | 0 | 24 | 9 | 6 | 0 | 39 | 2007 | 0 | 7 | 0 | 34 | 0 | 41 | 1 | 42 |
| 1951 | 0 | 27 | 7 | 6 | 0 | 40 | 2008 | 0 | 6 | 0 | 31 | 0 | 37 | 2 | 40 |
| 1952 | 0 | 40 | 2 | 6 | 0 | 48 | 2009 | 0 | 5 | 0 | 46 | 0 | 51 | 0 | 51 |
| 1953 | 0 | 99 | 2 | 4 | 0 | 105 | 2010 | 0 | 8 | 0 | 35 | 0 | 43 | 0 | 43 |
| 1954 | 0 | 116 | 1 | 7 | 0 | 124 | 2011 | 0 | 5 | 0 | 22 | 0 | 27 | 5 | 22 |
| 1955 | 0 | 66 | 1 | 9 | 0 | 76 | 2012 | 0 | 7 | 0 | 38 | 0 | 45 | 2 | 43 |
| 1956 | 0 | 161 | 1 | 6 | 0 | 168 | 2013 | 0 | 3 | 0 | 30 | 0 | 33 | 0 | 33 |
| 1957 | 0 | 94 | 1 | 4 | 0 | 99 | 2014 | 0 | 3 | 0 | 24 | 0 | 26 | 1 | 26 |
| 1958 | 0 | 163 | 2 | 6 | 0 | 171 | 2015 | 0 | 12 | 9 | 31 | 0 | 52 | 0 | 52 |
| 1959 | 0 | 327 | 1 | 8 | 0 | 336 | 2016 | 0 | 6 | 0 | 36 | 0 | 42 | 0 | 42 |
| 1960 | 0 | 80 | 1 | 9 | 0 | 90 | 2017 | 0 | 5 | 0 | 22 | 0 | 27 | 0 | 27 |
| 1961 | 0 | 110 | 1 | 12 | 0 | 123 | 2018 | 0 | 5 | 0 | 22 | 0 | 27 | 0 | 27 |
| 1962 | 0 | 100 | 0 | 8 | 0 | 108 | 2019 | 0 | 3 | 0 | 15 | $<1$ | 18 | 0 | 18 |
| 1963 | 0 | 172 | 0 | 19 | 0 | 191 | 2020 | 0 | 2 | 0 | 13 | 1 | 16 | 0 | 16 |
| 1964 | 0 | 159 | 1 | 24 | 0 | 184 | 2021 | 0 | 2 | 0 | 14 | 0 | 17 | 0 | 17 |



Figure 30.1. Landings of Sole in 7.bc as officially reported to ICES (1908-2021).




Figure 30.2. Sole in 7.b Irish otter trawl landings (top) effort (middle) and landings per unit of effort (LPUE; bottom) since 1995.

# 31 Sole (Solea solea) in Division 7.a (Irish Sea) 

## Type of assessment in 2022

This assessment is an update assessment.

## ICES advice applicable to 2022

ICES advises that when the MSY approach is applied, catches in 2022 should be no more than 787 tonnes.

## Comments made by the audit of last year's assessment

No major deficiencies for the sole assessment in the Irish Sea were reported.

### 31.1 General

## Stock description and management units

The sole fisheries in the Irish Sea are managed by TAC (see text tables below) and technical measures, with the assessment area corresponding to the stock area. Technical measures in force are minimum mesh sizes and minimum landing size ( 24 cm ). In addition, beam trawlers, fishing with mesh sizes equal to or greater than 80 mm , are obliged to have 180 mm mesh sizes in the entire upper half of the anterior part of their net. More details can be found in Council Regulation (EC) $\mathrm{N}^{\circ} 254 / 2002$ and the Stock Annex.

Since 2000, a spawning closure for cod has been in force. The first year of the regulation the closure covered the western and eastern Irish Sea. Since then, closure has been mainly in the western part whereas the sole fishery takes place mainly in the eastern part of the Irish Sea (Liverpool Bay and Cardigan Bay). No direct impact on the sole stock is expected from this closure.

For 2009, Council Regulation (EC) ${ }^{\circ} 43 / 2009$ allocates different amounts of $\mathrm{kW}{ }^{*}$ days by Member State and area to different effort groups of vessels depending on gear and mesh size. The area's are Kattegat, part of 3.a not covered by Skaggerak and Kattegat, ICES zone IV, EC waters of ICES zone 2.a, ICES zone 7.d, ICES zone 7.a, ICES zone $6 . a$ and EC waters of ICES zone 5.b. The grouping of fishing gear concerned are: bottom trawls, Danish seines and similar gear, excluding beam trawls of mesh size: TR1 ( $\geq 100 \mathrm{~mm}$ ) - TR2 ( $\geq 70$ and $<100 \mathrm{~mm}$ ) - TR3 ( $\geq 16$ and $<32 \mathrm{~mm}$ ); beam trawl of mesh size: BT1 ( $\geq 120 \mathrm{~mm}$ ) - BT2 ( $\geq 80$ and $<120 \mathrm{~mm}$ ); gillnets excluding trammelnets: GN1; trammelnets: GT1 and longlines: LL1.
For 2010-2016, Council Regulation (EC) N ${ }^{\circ} 53 / 2010$, Council Regulation (EC) N ${ }^{\circ} 57 / 2011$, Council Regulation (EC) $\mathrm{N}^{\circ} 43 / 2012$, Council Regulation (EC) $\mathrm{N}^{\circ} 40 / 2013$, Council Regulation (EC) $\mathrm{N}^{\circ} 43 / 2014$, Council Regulation (EC) $\mathrm{N}^{\circ} 2015 / 104$ and Council Regulation (EC) N ${ }^{\circ} 2016 / 72$ were updates of the Council Regulation (EC) N ${ }^{\circ} 43 / 2009$ with new allocations, based on the same effort groups of vessels and areas as stipulated in Council Regulation (EC) N43/2009.

Since the 1st of April 2015 all Belgian beam trawl vessels with mesh size of 80-119 mm fishing in ICES Division 7.a are obliged by national decree to use the 'Flemish Panel' to increase selectivity. This means the last tapered netting section of a beam trawl anterior is directly attached to the codend, the upper and lower netting sections are constructed of at least 120 mm mesh (as measured between the knots) and the stretched length is at least 3 m .

## Management applicable to 2021 and 2022

The TAC and the national quotas by country for 2021

| Species:Common sole <br> Solea solea |  | Zone:7 a <br> (SOL/07A.) |  |
| :--- | :--- | :--- | :--- | :--- |
| Belgium | 356 | Analytical TAC |  |
| France | 5 | Article 3 of Regulation (EC) No 847/96 shall not apply |  |
| Ireland | 104 | Article 4 of Regulation (EC) No 847/96 shall not apply |  |
| The Netherlands | 113 |  |  |
| Union | 578 |  |  |
| United Kingdom | 176 |  |  |
|  |  |  |  |
| TAC | 768 |  |  |

The TAC and the national quotas by country for 2022

| Species: | Common sole <br> Solea solea | Zone:7a <br> (SOL/07A.) |
| :--- | :--- | :--- |
| Belgium | 364 | Analytical TAC <br> Article 3 of Regulation (EC) No $847 / 96$ shall not <br> apply |
| Ireland | 5 | Article 4 of Regulation (EC) No 847/96 shall not <br> apply |
| Netherlands | 105 | 590 |
| Union | 181 |  |
| United Kingdom | 787 |  |

## Fishery in 2021

A full description of the fishery is provided in the Stock Annex, Section A2.
An overview of the landings data provided and used by the Working Group (WG) is shown in Table 31.1. The landings reached a level of 2808 t in the mid-1980s due to good recruitments in 1982-1984, but then subsequently dropped to a lowest of 818 t in 2000 . After a small increase to 1090 t in the beginning of the 2000s, the landings have fallen to under 350 t in 2008-2012. From 2013 to 2018 the landings continued to decrease as they dropped to under 150 t . In 2017, the record low value of 34 t was recorded. From 2016 to 2018, there has been no targeted fisheries for sole in ICES Division 7.a. Afterwards the landings increased again to about 400 t in 2019 and 2020 and 629 t in 2021.

In 2021, the WG estimated landings are 629 t , of which Belgium landed 75\% (473 t), Ireland 13\% ( 81 t ), $9 \%(56 \mathrm{t}$ ) by the UK (England and Wales) and the remainder by Northern Ireland, Scotland, Isle of Man and France. This corresponds to an international uptake of $82 \%$ of the agreed TAC in 2021 (768 t) and last year's forecast.

The WG estimate of the 2020 landings was not revised.

In $2021,87 \%$ of the landings were taken by beam trawls, $12 \%$ by otter trawls and $<1 \%$ by other gears.

### 31.2 Data

## Landings

Age compositions for 2021 were available from the countries that take the major part of the international landings (97\%) (Belgium, UK (E\&W) and Ireland). The raw age data were combined for the three countries without weighting. The combined ALK was applied to the raised length distribution of the national catches to obtain a combined age distribution. This distribution was applied to the landings from Northern Ireland, Scotland, Isle of Man and France to obtain the landings numbers-at-age for 2021 (Table 31.2, Figure 31.1). The standardised landings propor-tion-at-age is presented in Figure 31.2. Annual length distributions of the three major countries involved are given in Table 31.3. Because of the substantial reduction of the TAC in 2014-2018, sampling levels in this period were also substantially reduced.

Landings weights-at-age for 2021 were taken from the combined age-weight key (Table 31.4).
Stock weights-at-age for 2000-2021 were derived from the mean catch weights by cohort interpolation to the first of January (Rivard weight calculator) (Table 31.5).

Further details on raising methods are given in the stock annex.
As last year, the combined age data (calculated outside InterCatch) as well as the landings from Northern Ireland, Scotland, Isle of Man and France were uploaded to InterCatch. It should be noted that the international age distribution is uploaded as " $\mathrm{BE}^{\prime}$ " as no international country code is available in InterCatch at present.

## Discards

The available discard information (Table 31.6 and Figure 31.3) suggests that discarding is not a major problem in the Irish Sea sole fishery. However, discards have increased recently from 3.5\% (average 2016-2018) to 12\% (average 2019-2021). Belgian beam trawl length distributions of retained and discarded catches of sole for 2021 (Figure 31.3) indicate that predominantly 2 and 3year old fish are discarded. In 2020 and 2021, no observer information from the UK and Irish fleet was available.

As an attempt, estimating an overall discard rate for the stock, individual discard estimates for 2019-2021 from the main métiers and countries (Belgium, Ireland and UK) were averaged to obtain an overall discard rate (Table 31.6b). In 2020 and 2021, only discard information from the Belgian beam trawl fleet was provided. The percent of the métiers with discard information covering the total international landings is $90 \%, 74 \%$ and $75 \%$ for 2019,2020 and 2021 respectively. Assuming that discard rates do not change from the average of the last three years (2019-2021) and a fixed proportion of discards survive, a discard rate of around $12 \%$ (of the catch) could be assumed for this stock at the moment.

## Biological

Natural mortality, maturity and proportions of natural mortality and fishing mortality before spawning were set as in previous years, details of which can be found in the Stock Annex section B2.

## Surveys

The UK (E\&W) September beam-trawl survey (UK(E\&W)-BTS-Q3) was unable to cover the 7.a Division in 2020 due to the Covid-19 disruption. Therefore, the 2020 information is missing from the LPUE series (1988-2019,2021) (Table 31.7b and Figure 31.4). The UK (E\&W) March beamtrawl survey (UK(E\&W)-BTS-Q1) provides information from 1993 to 1998. From 2006 until 2010 the two UK beam trawl surveys have been used as tuning indices in the Irish Sea sole assessments (Table 31.8). Following the outcome of WKFLAT 2011, the March survey (UK(E\&W)-BTS-Q1) was omitted from the following assessments.

The LPUE from the UK(E\&W)-BTS-Q3 has fluctuated since the beginning of the time-series (1988) between 90 and $200 \mathrm{~kg} / 100 \mathrm{Km}$ fished. Since 2000, it has dropped gradually to the lowest value in 2012 ( $26.47 \mathrm{~kg} / 100 \mathrm{Km}$ fished). Thereafter, it gradually increased to $118.66 \mathrm{~kg} / 100 \mathrm{Km}$ fished in 2019. For 2021 a lower value of $63.6 \mathrm{~kg} / 100 \mathrm{Km}$ fished was noted.

The UK(E\&W)-BTS-Q3 survey was unable to cover the 7.a Division in 2020 due to Covid-19. The assessment was performed without tuning data for 2020.

Detailed information on the survey protocols and area coverage can be found in the Stock Annex.

## Commercial LPUE

Trends in LPUE and effort are given in Table 31.7 and Figures 31.5 and 31.6.
Commercial LPUE and effort data were available for Belgian beam trawlers, UK (E\&W) beam and otter trawlers and Irish otter and beam trawlers. It should be noted that the most recent LPUE values of the UK (E\&W) beam trawlers (2013-2021) and the UK (E\&W) otter trawlers (2014-2021) are based on days fished instead of hours fished. In 2013, the UK administration switched to the EU electronic logbook system. Therefore, a lot of the reported effort is missing and the 2013 value cannot be used as an absolute number. Details of the 2013 UK beam trawl were unavailable due to reduced numbers of trips reporting this gear specific effort information via the newly introduced e-logbook system. The otter trawl fleet effort reporting was unaffected by this as these vessels were not reporting their landings via this method in 2013. However, from 2014 onwards, both the UK beam trawl and otter trawl effort values (hours fished) are unavailable because of the reporting issues.

Effort from both Belgian and UK commercial beam trawl fleets increased from the early seventies until the beginning of the nineties. Since then UK beam trawl effort has shown a continuing declining trend. Inspection of an alternate effort indicator (days fished) suggests that the declining trend continues in the period 2013-2018, followed by a slight increase in 2019. In 2020 and 2021, effort continues to increase to a similar level as observed in 2007. In contrast, the Belgian beam trawl effort has shown a more fluctuating pattern. After the decline in the early nineties, it reached its highest level in 2002 and decreased again afterwards. For the period 2008-2012, it remained stable at a very low level but in 2013 it continued to decrease and in 2016, it dropped to the lowest level in the time-series. In 2017-2018, there's a slight increase. In the period 20192021, effort further increased to the level recorded in 2012. The substantial decrease of the Belgian and UK commercial beam trawl effort in the period 2013-2018, is in line with the substantial reductions of the TAC. From 2019 onwards, a sole-directed fisheries is again allowed and a higher TAC is set. This is clearly reflected in the higher activity of the Belgian beam trawlers from 2019 onwards.

The effort of the Irish beam trawlers shows a slow decline since 2004 and reached the lowest level in the time-series in 2013. Since 2014, the Irish beam trawl effort has increased, followed by
a decrease in 2020 and 2021. In 2008, all beam trawl fleets showed a substantial reduction in effort compared to 2007.

The effort from the UK otter trawlers remained stable until the beginning of the nineties. Since then the UK otter trawl effort has continuously declined and is at the lowest level in 2013. As, in 2015 and 2016 all otter trawl vessels active in the Irish Sea were under 12 m , no effort (days fished) was recorded. Since 2017, the otter trawl effort (days fished) fluctuates at a low level. The Irish otter trawlers have shown a striking reduction in effort since 2000, followed by a slight increase in the period 2010-2012. In 2017, the Irish otter trawl effort fell back to the lowest observed level in the time-series. It remains at this lower level, except for the slightly higher value observed in 2019.

LPUE for both UK and Belgian beam trawlers was at a high level in the late seventies and early eighties but since early 2000s, LPUE for these fleets has fluctuated at a lower level. In the period 2007-2009 there has been a small increase in the UK beam trawl LPUE. However, in 2012 the LPUE has dropped to a remarkable low level in the time-series ( $4.3 \mathrm{~kg} / \mathrm{hour}$ fished). An update for 2013-2017 was not available. However, the alternate LPUE indicator ( $\mathrm{kg} / \mathrm{days}$ fished) suggests that the UK beam trawl LPUE increased in 2015. For 2016-2018 no catches of sole and/or no effort were recorded therefore the LPUE is zero. After a slight increase in 2019, the LPUE further increased in 2020-2021 to a similar level as observed in 2009. The Belgian beam trawlers hold on to a higher LPUE value (18-20 kg/hour fished) for the period 2008-2012. However, in 2013 the LPUE decreased ( $13.2 \mathrm{~kg} /$ hour fished) and in 2017 it dropped to the lowest level in the time-series ( $3.8 \mathrm{~kg} /$ hour fished). In 2018, there's a slight increase to $5.4 \mathrm{~kg} /$ hour fished, followed by a substantial increase to $32.3 \mathrm{~kg} /$ hour fished in 2019 and $30.8 \mathrm{~kg} / \mathrm{hour}$ fished in 2020. In 2021 the LPUE further increased to the highest level of the time-series ( $46.17 \mathrm{~kg} / \mathrm{hour}$ ). The Irish beam trawl LPUE shows a gradually diminishing trend over the whole time-series. After the slight increase in 2013, it fell back to a record low level in 2016-2018. Since 2019 there's a slight increasing trend.

The UK otter trawl LPUE remained stable until the beginning of the 2000s but is at the record low level in 2012. The alternative LPUE indicator (kg/days fished) suggests that the declining trend continues after 2012. After the record low level in 2017 ( $5.6 \mathrm{~kg} /$ days fished), the LPUE gradually increases to 192.7 kg /days fished) in 2021. In 2012-2016, the LPUE of Irish otter trawlers is fluctuating at a lower level. In 2017-2021 a higher value was recorded.

In 2020 during which the COVID-19 disruptions took place, a shift between the UK beam and otter trawl fleet was noted, as the activity of the beam trawlers substantially increased whereas that of the otter trawlers was substantially reduced. Further, no substantial changes in effort or LPUE compared to 2019 were recorded.

## Historical Stock Development

In 2010, the Irish Sea sole assessment was based on XSA with two survey tuning indices (UK(E\&W)-BTS-Q3 and UK(E\&W)-BTS-Q1 (Table 31.8). The UK(E\&W)-BTS-Q1 indices only provide information for the years 1993 up to 1999 and therefore no longer contribute to the final survivor estimates. At WKFLAT 2011, the exclusion of the UK(E\&W)-BTS-Q1 from the assessment was investigated and it was found that there was little effect on the catchability residuals and that the retrospective pattern was slightly improved. WKFLAT 2011 therefore decided to omit this survey from the assessment. In 2020 and 2021, the assessment was performed without tuning data for 2020, as the UK(E\&W)-BTS-Q3 could not take place in Division 7.a due to the Covid-19 disruptions.

### 31.3 Stock assessment

## Data screening

The age range for the analysis was 2-8+.
The screening of the tuning indices (UK(E\&W)-BTS-Q3) showed good cohort tracking (Figure 31.7) and consistency between ages for year-class strength (Figure 31.8).

## Final Update Assessment

The model settings for the final assessment are summarized below

| Assmnt Year | $: 2010$ | $: 2011-2021$ |
| :--- | :--- | :--- |
| Assmnt Model | $:$ XSA | $:$ XSA |
| Fleets | $:$ | $:$ |
| Bel Beam Trwl | $:$ omitted | :omitted |
| UK Trawl | $:$ omitted | :omitted |
| UK Sept BTS | $: 1988-2009[2-7]$ | $: 1988-$ assessment year-1 ${ }^{*}$ [2-7] |
| UK Mar BTS | $: 1993-1999[2-7]$ | :omitted |
| Time Ser. Wts | $:$ linear 20 yrs | :no taper weighting |
| Power Model | $:$ none | :none |
| Q plateau | $: 7$ | $: 4$ |
| Shk se | $: 1.5$ | $: 1.5$ |
| Shk age-yr | $: 5$ yrs 3 ages | $: 5$ yrs 3 ages |
| Pop Shk se | $: 0.3$ | $: 0.3$ |
| Prior Wting | $:$ none | $:$ none |
| Plusgroup | $: 8$ | $: 8$ |
| Fbar | $: 4-7$ | $: 4-7$ |

*2020 not available due to COVID-19 and therefore not included in the assessment.

The final XSA output is given in Table 31.9 (diagnostics), Table 31.10 (fishing mortalities) and Table 31.11 (stock numbers). Log catchability residuals for the final assessment are given in Figure 31.9. A summary of the XSA results is given in Table 31.12 and trends in yield, fishing mortality, recruitment and spawning-stock biomass are shown in Figure 31.10. Retrospective patterns for the final run are shown in Figure 31.11.

The UK(E\&W)-BTS-Q3 survey was unable to cover Division 7.a in 2020 due to COVID-19. Last year's and this year's assessment were performed without tuning data for 2020. Last year, the recruitment estimate for 2020 was considered uncertain and was replaced by the geometric mean of recruitment (GM, 2015-2019). With the addition of the 2021 catch data and the 2021 UK(E\&W)-BTS-Q3 tuning information, the 2020 recruitment ( 7051 thousand fish) is estimated to be higher than the geometric mean of recruitment (2931 thousand fish, GM, 2015-2019). Both the age 3 catch numbers in 2021 (Figure 31.1) and the age 3 UK(E\&W)-BTS-Q3 index of 2021 indicate that 2018 is a strong year class.

Adding the 2021 data to the time-series did not cause any additional anomalies compared to last year. The log catchability residual pattern showed no trends apart from the year effect in 2016 and 2021. The positive residuals (higher estimates from the UK(E\&W)-BTS-Q3 fleet compared to
the VPA estimates) in 2016 are likely due to the fact that de age composition in the catch is flattened. For 2021 all negative residuals (lower estimates from the UK(E\&W)-BTS-Q3 fleet compared to the VPA estimates) were noted.

The survivor estimates and fishing mortality estimates are almost entirely determined by the UK(E\&W)-BTS-Q3 survey as it gets a high weighting (>96\%) at all ages.

A Mohn's rho analysis was conducted based on the XSA stock assessment results, i.e. the last data year (2021) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 4-7) | recruitment |
| :--- | :--- | :--- | :--- |
| Mohn's rho value | 0.086 | -0.081 | -0.085 |

The Mohn's rho values for this assessment are low and are well within the ICES WKFORBIAS thresholds ( $+0.20,-0.15$ ), i.e. the current assessment indicates sufficient consistency for advice purposes. There is a slight tendency to overestimate the SSB and to underestimate the fishing mortality.

## Comparison with previous assessments

A comparison of the estimates of this year's assessment with last year's is given in Figure 31.12. Trends in fishing mortality, SSB and recruitment are very similar.

The geometric mean of recruitment (2931 thousand fish, GM, 2015-2019) that replaced the 2020 XSA recruitment estimate in last year's assessment, was revised upwards by $141 \%$ ( 7051 thousand fish) in this year's assessment. The 2019 and 2018 recruitments were revised downwards by $14 \%$ and $16 \%$ respectively in this year's assessment. In last year's assessment, F and SSB for 2020 were estimated to be 0.131 and 3493 t respectively; this year's estimates for 2020 are 0.149 and 3298 t , an upward revision of $14 \%$ for F and a downward revision of $6 \%$ for SSB.

## State of the stock

Estimated trends of Irish Sea sole landings, SSB, fishing mortality and recruitment are presented in Table 31.12 and Figure 31.12. Since the late eighties the landings of Irish Sea sole have been declining to the lowest level of the time-series (34t) in 2017. SSB has been at a higher level until the late eighties. Since then SSB has been fluctuating between $B_{p a}$ and Blim and since 2004 it dropped below $\mathrm{B}_{\mathrm{lim}}$. After the record low value in 2014 (866 t), SSB gradually increased but remains below just below MSYB trigger $^{\text {in }} 2021$ ( 3385 t ). High fishing mortalities were observed during the late eighties until the mid-nineties. Thereafter fishing mortality declined to a level fluctuating just above Flim. From 2013 onwards, fishing mortality has dropped under the level of $\mathrm{F}_{\mathrm{pa}} / \mathrm{F}_{\mathrm{ms}}$. In 2018, the lowest level of the time-series was recorded (0.015). The decline in F is supported by a substantial reduction of the TAC in this period. As the TAC increased and the sole targeted fisheries was again permitted, F increased to 0.156 in 2019 and 0.149 in 2020. In 2021 F increased to above $\mathrm{F}_{\mathrm{pa}} / \mathrm{F}_{\text {mSY }}$. Since 2001 recruitment has been well below the mean ( 5418 thousand fish) and the 2011 recruitment (year class 2009) is estimated to be the lowest in the time-series (633 thousand fish). The 2016 recruitment ( 3931 thousand fish, year class 2014) is estimated to be six times higher than the record low recruitment in 2011. Thereafter, higher and lower recruitments alternate. The 2018 year class ( 7051 thousand 2 year old fish) is estimated to be highest
recorded in the last 20 years, followed by the weaker 2019 year class ( 1405 thousand 2 year old fish).

### 31.4 Short-term projections

## Estimating year-class abundance

The 2011-2014 recruitments have been the lowest in the time-series. Higher recruitment was observed from 2015 onwards. However, the 2019 year class ((2021 recruitment) is now estimated at 1405 thousand fish at age 2 (Table 31.12), which is $52 \%$ lower than the short term GM (2015-2019) (2931 thousand fish) used in last year's forecast.

The age 2 estimates are almost solely coming from the UK(E\&W)-BTS-Q3. From 2010 to 2014, the UK(E\&W)-BTS-Q3 abundance index for age 2 fluctuated around the level of the lowest abundance in 2011 (0.29). In 2016 (2.97) and 2018 (2.18) again higher age 2 abundance indices were noted. The 2021 UK(E\&W)-BTS-Q3 abundance index for age $2(0.74)$ is around the same level of 2017 (0.8).

## Forecast assumptions

Figure 31.13 shows three different targets for the intermediate year: $\mathrm{F}_{-}$last ( $\mathrm{F}=\mathrm{F}_{2021}$ or status quo), F_average ( $\mathrm{F}=\mathrm{F}_{\text {average 2019-2021 }}$ ), and TAC. F estimates decreased slightly in 2020 compared to 2019 but increased in 2021.

The F in 2021 ( 0.259 ) is higher than in previous years ( $\mathrm{F}_{2020}=0.149, \mathrm{~F}_{2019}=0.156$ ) and the assumed 2022 landings using a status quo fishing mortality scaled to 2021 ( 815 t ) are higher than the 2022 TAC ( 787 t ). Landings and catch (discards calculated based on an overall discard rate) have been below the international TAC for 2021 (Figure 31.14), whereas in previous years catch and/or landings were closer to the TAC. This could possibly be allocated to the Brexit and the fact that the 2021 TAC was available at a later stage during the year. As this was a temporary issue, it seemed reasonable that the landings in 2022 would be in line with the ICES advice/TAC. Therefore, the working group agreed to use a landings constraint ( 696 t ) for the intermediate year (2022) like previously assumed for this stock.

As input for the forecast fishing mortality was calculated as the mean of 2019-2021, scaled to 2021 (0.259). Catch and stock weights-at-age were also averages for the years 2019-2021. Population numbers at the start of 2022 for ages 3 and older, were taken from the XSA output. The Stock Annex recommends using a short-term geometric mean (assessment year minus 10 up to assessment year minus 2) for the recruitment in the short-term forecast.

The forecast was conducted with FLR's FLash R package using the output from the landings only XSA assessment. The resulting yield was obtained by adding discards to the landing with an average discard rate of the last three historical years (2019-2021, 0.12\%).
The input for the short-term catch predictions and sensitivity analysis is given in Table 31.13.

## MSY forecast

As the SSB in 2023 ( 3299 t ) is assumed to be below MSYB ${ }_{\text {trigger, }}$ the Fmsy target (0.2) is rescaled to 0.189 (FMSY $\times$ SSB2023/MSYB trigger ).

Table 31.15 and Figure 31.15 show a detailed output of the forecast targeting $\mathrm{F}=\mathrm{F}_{\text {msy }} \times$ SSB2023/MSYB trigger for 2023-2024 and Figure 31.16 shows the year classes contributing to the
forecast yield and SSB. The assumed short-term GM recruitments accounts for about $10 \%$ of the landings in 2023 and about $21 \%$ of the 2024 SSB.

Implementing the MSY approach with F $=\mathrm{F}_{\text {MSY }} \times \mathrm{SSB}_{2023 / M S Y B_{\text {trigger }}}=0.189$, leads to a total yield of 605 t in 2023, and an SSB of 31291 t in 2024.

The advice is a reduction of $23 \%$ compared to the advice for 2022 . This decline is because of (1) a projected decline in the SSB caused by low recruitment in 2021 and (2) a downward revision of stock size caused by lower 2021 survey indices for age $4+$ and lower mean stock weights-at-age.

## Additional options

A management options table is provided in Table 31.14

### 31.5 Biological reference points

The most recent reference points for this stock were developed by WKMSYREF4 in 2015 (ICES, 2016). These reference points are presented in the following table. Please note that ICES changed the basis for $\mathrm{F}_{\mathrm{pa}}$ to $\mathrm{F}_{\mathrm{p} .05}$ in 2021, and the updated $\mathrm{F}_{\mathrm{pa}}$ value is shown here.

| Reference points | ACFM 2007 onwards | 2016 onwards | 2021 onwards |
| :---: | :---: | :---: | :---: |
| $\mathrm{F}_{\text {MSY }}$ | 0.16 (PLOTMSY, WG2010) | 0.20 (Eqsim, WKMSYREF 4) | 0.20 (Eqsim, WKMSYREF 4) |
| $\mathrm{F}_{\text {lim }}$ | 0.4 (based on $\mathrm{F}_{\text {loss }}$ ) | 0.29 (based on simulated recruitment to give median biomass $=\mathrm{B}_{\text {lim }}$ ) | 0.29 (based on simulated recruitment to give median biomass $=B_{\text {lim }}$ ) |
| FPA | 0.3 (high probability of avoiding $\mathrm{F}_{\text {lim }}$ ) | 0.21 ( $\mathrm{Flim}^{*} 1.4$ ) | 0.22 ( $F_{\text {p. } 05} ; F$ that leads to $S S B \geq B_{\text {lim }}$ with $95 \%$ probability) |
| $\mathrm{Blim}_{\text {lim }}$ | 2200 t ( $\mathrm{B}_{\text {loss }}$ estimated in 2007) | 2500 t (lowest value with above average recruitment) | 2500 t (lowest value with above average recruitment) |
| $\mathrm{B}_{\text {PA }}$ | $3100 \mathrm{t}\left(\mathrm{B}_{\mathrm{pa}} \sim \mathrm{B}_{\lim } * 1.4\right)$ | 3500 t ( Blim *1.4) | 3500 t ( Blim *1.4) |
| $B_{\text {trigger }}$ | $\mathrm{B}_{\text {PA }}$ | 3500 t | 3500 t |

### 31.6 Management plans

No management plan is currently in place for Irish Sea sole.

### 31.7 Uncertainties and bias in assessment and forecast

## Sampling

The deteriorating quality of the historic catch numbers-at-age data was considered to be a consequence of the low biological sampling intensity, and in particular the limited sampling in the first quarter. Therefore, the combined age distribution was introduced in 2000 as an alternative method for raising the international catch numbers-at-age. The mean catch weights from this combined key were taken and the stock weights-at-age were obtained using a cohort interpolation method from the catch weights-at-age. Under the DCF there is an initiative to co-ordinate
sampling across the three countries involved in the fishery. However, as the TAC is substantially reduced in recent years, sampling levels were also significantly reduced. In 2019 the TAC increased again and additionally a scientific sole quota is reserved for Belgian vessels fishing in ICES Division 7.a to assure a qualitative sampling. Due to Covid-19, only discard information from the Belgian beam trawl fleet was provided for 2020 and 2021.

## Landings

There is no reliable information on the accuracy of the landing statistics. For the period 20052012, the total TAC uptake was only in the range of $50-98 \%$. In this context, misreporting was not considered to be a major problem. In the most recent years, the TAC was substantially reduced and was restrictive in 2013 and 2014. In 2015-2020, $84 \%-97 \%$ of the TAC has been taken. 629 t sole were landed and 61 t were discarded in 2021, a total catch of 690 t , while the agreed TAC was 787 t .

## Discards

The absence of discard data in the assessment is considered to have a minor effect on the quality of the assessment as the average discarding by weight has been low in the past (3-8\%). However, higher discard rates were recorded in 2019 ( $14 \%$ ) and $2020(12 \%)$. The most recent discard information indicates a decrease in 2021 (9\%). It might be recommended to include discards in the next benchmark assessment.

## Effort

There are no indications of Irish Sea sole fisheries misreporting effort. Effort in beam trawl fisheries that target sole has declined substantially in the last few years in accordance with the significant reductions in TAC. In 2019-2021 higher effort values were recorded as the TAC increased and sole directed fisheries were again allowed.

## Surveys

The UK(E\&W)-BTS-Q3 survey appears to track year-class strength well. As previously investigated, this tuning fleet is also consistent in estimating year-class strength of the same year class at different ages. Therefore, the Working Group had confidence in using the UK(E\&W)-BTS-Q3 survey as the only tuning fleet. The bias problem in the assessment maybe the result of the precise survey and less precise catch-at-age data. The UK(E\&W)-BTS-Q3 survey data for 2020 were not available due to COVID-19.

## Model formulation

At present XSA is used to assess Irish Sea sole. In the WG of 2007 the model settings were changed which had a considerable impact on the estimates of SSB and fishing mortality. Due to these major revisions, ACFM changed the biomass reference points at its meeting of 2007. In the next two update assessments (2008-2009) no major changes were apparent. In the assessment of 2011, the settings were changed according to the outcome of WKFLAT 2011. The following assessments were update assessments.

### 31.8 Recommendations for next Benchmark

The assessment diagnostics indicate a good correlation between the catch data and the survey tuning series. However, in the recent years there has been great uncertainty from the fishing industry on the actual status of the sole stock in the Irish Sea. Fishermen are concerned that due to ecosystem changes and the changing fishing behaviour in the Irish Sea, science is no longer capturing the current situation. Because of this mismatch, an EU action plan for the Irish Sea fisheries was set up. First, a comparative fishing study was suggested to compare the catch efficiency between the UK-BTS-Q3 and a Belgian commercial vessel. Secondly, a pilot industryscience beam trawl survey should reveal the spatial distribution of sole. The outcome of those work packages will indicate whether the data gathered by the UK-BTS-Q3 is still representative for the current situation or whether the implementation of an additional (annual) industry-science industry survey is needed. Thirdly, stock identification techniques (i.e. genetic fingerprinting and otolith shape analysis) will be performed to give insight on the origin and potential migration routes of sole that is caught in the Irish Sea.

The industry survey was not able to identify other areas of importance for sole in the Irish Sea than is already covered by the UK-BTS-Q3. Also, catchability and composition of catches in both surveys were comparable. These results suggest that the UK-BTS-Q3 gives a good representation of sole abundance and that an annual industry survey additional to this survey would not be of added value to the assessment. With regards to the stock identification study, the combination of otolith shape analysis and genetic markers (SNPs) show subtle differences between the Irish Sea, Celtic Sea and Bristol Channel populations. However more samples from the different areas and from different years need to be analysed to reveal what is driving these differences. Also, in the attempt to effectively reassign adult sole to their place of origin, it would be preferable to include a third stock identification technique: micro-chemical fingerprinting. Despite many questions yet unsolved, the pilot industry survey delivered valuable information that can be added to an ecosystem model for the Irish Sea (one of the aims of WKIRISH: an ecosystem benchmark for the Irish Sea). Moreover, the survey was an example of a fruitful cooperation between fishermen and fisheries scientists and gave useful insights on how to cooperate with the fishing industry and to gain their trust in the collection of fisheries-independent data.

A future benchmark might look into including discard estimates in the assessment and estimating historical discards. An alternative assessment model such as a state-space stock assessment model (SAM) would be beneficial and allows to propagate the main uncertainties into the forecasts properly.

### 31.9 Management considerations

There is a stock-recruitment relationship for this stock and evidence of reduced recruitment at low levels of SSB. However, the recruitment for higher levels of SSB is less well defined (Figure 31.17).

Recruitment at age 2 has been well below average since 2001. In 2016 and 2018, recruitment is estimated to be six times higher than the record low levels in 2011-2014. SSB has increased since 2014 and is between $B_{l i m}$ and MSY Btrigger in 2019, 2020 and 2021. SSB in 2021 was 3385 and slightly below MSY Btrigger ( 3500 t ). Over the last couple of years fishing mortality has decreased to close to zero. In 2019, fishing mortality started to increase and is above FMSY in 2021.

Sole is caught in a mixed-fishery with other flatfish as well as gadoids. Information from observer trips indicates that discarding of sole is relatively low.

### 31.10 Ecosystem considerations

Sole and plaice are primarily targeted by beam trawl fisheries. Beam trawling, is known to have an impact on the benthic communities, although less so on soft substrates and in areas which have been historically exploited by this fishing method. Some beam trawlers are using benthic drop-out panels that release about $75 \%$ of benthic invertebrates from the catches. Full square mesh codends are being tested in order to reduce the capture of benthos further and improve the selection profile of gadoids (Connolly, P.L. et al., 2009).

A complete ecosystem overview can be found in the stock annex Section A.3.

### 31.11 References

Connolly, P.L., Kelly, E., Dransfeld, L., Slattery, N., Paramor, O.A.L., and Frid, C.L.J. 2009. MEFEPO North Western Waters Atlas. Marine Institute.

ICES. 2014a. Report of the Workshop to consider reference points for all stocks (WKMSYREF2), 8-10 January 2014, ICES Headquarters, Copenhagen, Denmark. ICES CM 2014/ACOM:47. 91 pp.
ICES. 2014b. Report of the Joint ICES-MYFISH Workshop to consider the basis for FMSY ranges for all stocks (WKMSYREF3), 17-21 November 2014, Charlottenlund, Denmark. ICES CM 2014/ACOM:64. 147 pp.

ICES. 2016a. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 183 pp.

ICES. 2016b. EU request to ICES to provide FMSY ranges for selected stocks in ICES subareas 5 to 10, ICES special request advice. 5 February 2016 Version 2; 13 May 2016.

Table 31.1. Sol.27.7a - Nominal landings (tonnes) as officially reported by ICES, and working group estimates of the landings. Last year's landings are preliminary.

| Year |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \frac{\varepsilon}{3} \\ & \frac{10}{00} \\ & \frac{0}{0} \end{aligned}$ |  | $\begin{aligned} & \text { 들 } \\ & \text { N } \\ & \underline{\underline{0}} \end{aligned}$ |  | $\begin{aligned} & \underset{3}{3} \\ & \underset{\sim}{4} \\ & \text { y } \end{aligned}$ |  |  |  |  |  | Total used by WG | $\underset{\downarrow}{\text { U }}$ |
| 1973 | 793 | 12 | 27 | 281 | 258 | - | 46 | 11 | 1428 | 0 | 1428 |  |
| 1974 | 664 | 54 | 28 | 320 | 218 | - | 23 | - | 1307 | 0 | 1307 |  |
| 1975 | 805 | 59 | 24 | 234 | 281 | - | 24 | 15 | 1442 | -1 | 1441 |  |
| 1976 | 674 | 72 | 74 | 381 | 195 | - | 49 | 18 | 1463 | 0 | 1463 |  |
| 1977 | 566 | 39 | 84 | 227 | 160 | - | 49 | 21 | 1146 | 1 | 1147 |  |
| 1978 | 453 | 65 | 127 | 177 | 189 | - | 57 | 30 | 1098 | 8 | 1106 |  |
| 1979 | 779 | 48 | 134 | 247 | 290 | - | 47 | 42 | 1587 | 27 | 1614 |  |
| 1980 | 1002 | 41 | 229 | 169 | 367 | - | 44 | 68 | 1920 | 21 | 1941 |  |
| 1981 | 884 | 13 | 167 | 186 | 311 | - | 41 | 45 | 1647 | 20 | 1667 |  |
| 1982 | 669 | 9 | 161 | 138 | 277 | - | 31 | 44 | 1329 | 9 | 1338 |  |
| 1983 | 544 | 3 | 203 | 224 | 219 | - | 33 | 29 | 1255 | -86 | 1169 |  |
| 1984 | 425 | 10 | 187 | 113 | 230 | - | 38 | 17 | 1020 | 38 | 1058 |  |
| 1985 | 589 | 9 | 180 | 546 | 269 | - | 36 | 28 | 1657 | -511 | 1146 |  |
| 1986 | 930 | 17 | 235 | - | 637 | 1 | 50 | 46 | 1916 | 79 | 1995 |  |
| 1987 | 987 | 5 | 312 | - | 599 | 3 | 72 | 63 | 2041 | 767 | 2808 | 2100 |
| 1988 | 915 | 11 | 366 | - | 507 | 1 | 47 | 38 | 1885 | 114 | 1999 | 1750 |
| 1989 | 1010 | 5 | 155 | - | 613 | 2 | - | 38 | 1823 | 10 | 1833 | 1480 |
| 1990 | 786 | 2 | 170 | - | 569 | 10 | - | 39 | 1576 | 7 | 1583 | 1500 |
| 1991 | 371 | 3 | 198 | - | 581 | 44 | - | 26 | 1223 | -11 | 1212 | 1500 |
| 1992 | 531 | 11 | 164 | - | 477 | 14 | - | 37 | 1234 | 25 | 1259 | 1350 |
| 1993 | 495 | 8 | 98 | - | 338 | 4 | - | 28 | 971 | 52 | 1023 | 1000 |
| 1994 | 706 | 7 | 226 | - | 409 | 5 | - | 14 | 1367 | 7 | 1374 | 1500 |
| 1995 | 675 | 5 | 176 | - | 424 | 12 | - | 8 | 1300 | -34 | 1266 | 1300 |
| 1996 | 533 | 5 | 133 | 149 | 194 | 4 | - | 5 | 1023 | -21 | 1002 | 1000 |
| 1997 | 570 | 3 | 130 | 123 | 189 | 5 | - | 7 | 1027 | -24 | 1003 | 1000 |


| Year |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

${ }^{1} 1989$ onwards: N. Ireland included with England \& Wales.

Table 31.2. Sol.27.7a - Landings numbers-at-age (in thousands).

| year |  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 29 | 895 | 1009 | 467 | 1457 | 289 | 228 | 803 | 265 | 729 | 91 | 74 | 14 | 333 | 6683 |
| 1971 | 113 | 434 | 2096 | 1130 | 232 | 878 | 141 | 106 | 327 | 376 | 265 | 298 | 54 | 320 | 6769 |
| 1972 | 31 | 673 | 730 | 1538 | 537 | 172 | 522 | 97 | 46 | 279 | 142 | 152 | 98 | 164 | 5182 |
| 1973 | 368 | 363 | 2195 | 557 | 815 | 267 | 112 | 329 | 74 | 104 | 150 | 135 | 87 | 152 | 5708 |
| 1974 | 25 | 891 | 576 | 1713 | 383 | 422 | 232 | 58 | 226 | 44 | 55 | 103 | 110 | 143 | 4981 |
| 1975 | 262 | 733 | 2386 | 539 | 842 | 157 | 227 | 158 | 91 | 139 | 24 | 24 | 110 | 233 | 5924 |
| 1976 | 29 | 375 | 1331 | 2329 | 247 | 544 | 134 | 151 | 80 | 16 | 98 | 28 | 9 | 223 | 5594 |
| 1977 | 221 | 416 | 1292 | 774 | 1066 | 150 | 218 | 89 | 64 | 46 | 7 | 63 | 49 | 112 | 4565 |
| 1978 | 65 | 958 | 649 | 1009 | 442 | 638 | 98 | 204 | 29 | 69 | 33 | 16 | 48 | 90 | 4347 |
| 1979 | 108 | 1027 | 3432 | 829 | 637 | 326 | 285 | 65 | 76 | 20 | 65 | 6 | 1 | 102 | 6978 |
| 1980 | 187 | 940 | 1969 | 3057 | 521 | 512 | 361 | 352 | 45 | 107 | 53 | 26 | 14 | 187 | 8333 |
| 1981 | 70 | 580 | 1668 | 1480 | 1640 | 114 | 184 | 86 | 258 | 22 | 130 | 26 | 22 | 137 | 6418 |
| 1982 | 8 | 346 | 1241 | 1298 | 711 | 641 | 91 | 113 | 23 | 81 | 46 | 10 | 2 | 31 | 4643 |
| 1983 | 37 | 165 | 998 | 758 | 757 | 416 | 334 | 69 | 74 | 35 | 83 | 23 | 36 | 55 | 3839 |
| 1984 | 651 | 786 | 380 | 610 | 343 | 424 | 178 | 251 | 23 | 30 | 19 | 36 | 3 | 17 | 3750 |
| 1985 | 154 | 1600 | 1085 | 343 | 334 | 164 | 259 | 188 | 127 | 45 | 22 | 6 | 37 | 55 | 4418 |
| 1986 | 141 | 3334 | 3465 | 960 | 235 | 277 | 210 | 187 | 125 | 157 | 27 | 46 | 22 | 74 | 9260 |
| 1987 | 189 | 3347 | 4104 | 3184 | 844 | 307 | 224 | 139 | 153 | 87 | 87 | 17 | 17 | 84 | 12783 |
| 1988 | 32 | 444 | 4747 | 2100 | 1309 | 203 | 83 | 76 | 45 | 93 | 70 | 62 | 7 | 80 | 9350 |
| 1989 | 179 | 771 | 775 | 3979 | 1178 | 552 | 121 | 23 | 28 | 8 | 41 | 4 | 8 | 22 | 7689 |
| 1990 | 564 | 1185 | 986 | 598 | 2320 | 592 | 333 | 38 | 17 | 18 | 13 | 11 | 5 | 31 | 6713 |
| 1991 | 1316 | 1269 | 841 | 300 | 226 | 1172 | 255 | 125 | 27 | 4 | 6 | 14 | 5 | 23 | 5583 |
| 1992 | 363 | 2431 | 917 | 556 | 190 | 156 | 523 | 217 | 156 | 23 | 3 | 1 | 0 | 6 | 5541 |
| 1993 | 83 | 543 | 1965 | 559 | 251 | 199 | 147 | 257 | 114 | 93 | 19 | 12 | 10 | 34 | 4285 |
| 1994 | 122 | 1343 | 1070 | 1579 | 394 | 133 | 98 | 141 | 171 | 37 | 55 | 4 | 8 | 10 | 5166 |
| 1995 | 132 | 920 | 1444 | 737 | 1010 | 179 | 62 | 48 | 61 | 80 | 32 | 40 | 9 | 18 | 4771 |
| 1996 | 60 | 469 | 1188 | 741 | 430 | 509 | 142 | 49 | 28 | 37 | 35 | 23 | 14 | 19 | 3745 |
| 1997 | 790 | 714 | 475 | 711 | 409 | 258 | 295 | 85 | 58 | 34 | 13 | 26 | 5 | 15 | 3889 |
| 1998 | 167 | 1728 | 466 | 256 | 315 | 191 | 126 | 150 | 51 | 45 | 18 | 17 | 6 | 10 | 3546 |
| 1999 | 301 | 1069 | 1259 | 297 | 115 | 136 | 82 | 37 | 45 | 22 | 10 | 5 | 8 | 23 | 3410 |
| 2000 | 178 | 906 | 907 | 600 | 150 | 55 | 70 | 53 | 24 | 45 | 21 | 6 | 13 | 26 | 3053 |
| 2001 | 240 | 1438 | 822 | 717 | 511 | 80 | 65 | 67 | 58 | 28 | 20 | 8 | 6 | 20 | 4080 |
| 2002 | 148 | 930 | 1623 | 740 | 575 | 254 | 79 | 30 | 48 | 24 | 12 | 13 | 4 | 6 | 4486 |
| 2003 | 437 | 825 | 966 | 795 | 302 | 217 | 205 | 29 | 14 | 24 | 15 | 13 | 18 | 26 | 3888 |
| 2004 | 299 | 862 | 342 | 368 | 304 | 139 | 57 | 54 | 14 | 12 | 10 | 9 | 4 | 20 | 2495 |
| 2005 | 536 | 1052 | 626 | 271 | 314 | 279 | 141 | 75 | 77 | 13 | 7 | 18 | 15 | 22 | 3445 |
| 2006 | 112 | 670 | 649 | 203 | 113 | 151 | 133 | 86 | 59 | 41 | 14 | 5 | 9 | 31 | 2276 |
| 2007 | 171 | 356 | 348 | 243 | 86 | 41 | 63 | 68 | 53 | 25 | 34 | 10 | 5 | 40 | 1541 |


| year/age 2 | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5 +}$ | Total |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2008 | 99 | 353 | 190 | 195 | 156 | 56 | 31 | 45 | 42 | 35 | 8 | 24 | 5 | 20 | 1258 |
| 2009 | 92 | 414 | 333 | 146 | 132 | 127 | 14 | 18 | 37 | 26 | 25 | 12 | 13 | 17 | 1405 |
| 2010 | 22 | 336 | 233 | 177 | 65 | 72 | 69 | 19 | 8 | 14 | 12 | 16 | 4 | 16 | 1063 |
| 2011 | 17 | 225 | 401 | 176 | 97 | 54 | 50 | 31 | 9 | 5 | 6 | 5 | 4 | 12 | 1091 |
| 2012 | 17 | 148 | 311 | 274 | 116 | 52 | 26 | 32 | 29 | 3 | 4 | 5 | 5 | 11 | 1034 |
| 2013 | 23 | 99 | 75 | 106 | 78 | 34 | 32 | 10 | 12 | 11 | 6 | 2 | 3 | 6 | 497 |
| 2014 | 12 | 49 | 59 | 37 | 38 | 51 | 17 | 11 | 8 | 5 | 5 | 2 | 2 | 6 | 303 |
| 2015 | 15 | 36 | 37 | 30 | 17 | 21 | 35 | 11 | 7 | 5 | 4 | 4 | 1 | 7 | 231 |
| 2016 | 1 | 18 | 22 | 14 | 10 | 7 | 8 | 11 | 3 | 4 | 1 | 2 | 2 | 2 | 103 |
| 2017 | 2 | 41 | 19 | 15 | 5 | 6 | 2 | 3 | 4 | 1 | 1 | 1 | 0 | 1 | 102 |
| 2018 | 4 | 22 | 46 | 14 | 9 | 3 | 2 | 1 | 2 | 2 | 1 | 1 | 0 | 1 | 107 |
| 2019 | 48 | 553 | 279 | 300 | 89 | 64 | 20 | 14 | 11 | 13 | 9 | 8 | 3 | 4 | 1416 |
| 2020 | 96 | 282 | 425 | 176 | 192 | 79 | 44 | 26 | 9 | 8 | 10 | 13 | 7 | 13 | 1379 |
| 2021 | 18 | 602 | 410 | 437 | 202 | 292 | 69 | 52 | 23 | 18 | 12 | 13 | 11 | 27 | 2187 |

Table 31.3. Sol.27.7a - Annual length distributions by country (2021).
$\left.\begin{array}{lccc}\hline \text { UK (England } \& \text { Wales) } & \text { Belgium } \\ \text { Length (cm) } \\ \text { All gears }\end{array}\right)$

Table 31.4. Sol.27.7a - Landing weights-at-age (kg).

| year/age |  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 0.13 | 0.153 | 0.178 | 0.204 | 0.232 | 0.26 | 0.29 | 0.321 | 0.353 | 0.387 | 0.422 | 0.458 | 0.495 | 0.533 |
| 1971 | 0.152 | 0.178 | 0.204 | 0.23 | 0.257 | 0.284 | 0.312 | 0.34 | 0.369 | 0.398 | 0.427 | 0.457 | 0.487 | 0.517 |
| 1972 | 0.126 | 0.164 | 0.201 | 0.237 | 0.272 | 0.306 | 0.338 | 0.369 | 0.4 | 0.428 | 0.456 | 0.483 | 0.508 | 0.533 |
| 1973 | 0.151 | 0.178 | 0.204 | 0.23 | 0.256 | 0.283 | 0.309 | 0.335 | 0.361 | 0.387 | 0.413 | 0.439 | 0.464 | 0.49 |
| 1974 | 0.138 | 0.174 | 0.209 | 0.241 | 0.272 | 0.301 | 0.328 | 0.353 | 0.377 | 0.399 | 0.419 | 0.437 | 0.453 | 0.468 |
| 1975 | 0.13 | 0.172 | 0.21 | 0.244 | 0.275 | 0.303 | 0.327 | 0.347 | 0.364 | 0.378 | 0.387 | 0.394 | 0.396 | 0.396 |
| 1976 | 0.12 | 0.161 | 0.2 | 0.239 | 0.276 | 0.313 | 0.348 | 0.383 | 0.416 | 0.449 | 0.48 | 0.511 | 0.541 | 0.569 |
| 1977 | 0.085 | 0.146 | 0.202 | 0.251 | 0.293 | 0.33 | 0.36 | 0.384 | 0.401 | 0.413 | 0.418 | 0.417 | 0.409 | 0.395 |
| 1978 | 0.093 | 0.147 | 0.197 | 0.243 | 0.286 | 0.326 | 0.361 | 0.394 | 0.422 | 0.447 | 0.468 | 0.486 | 0.5 | 0.511 |
| 1979 | 0.134 | 0.165 | 0.199 | 0.234 | 0.271 | 0.311 | 0.352 | 0.395 | 0.441 | 0.488 | 0.537 | 0.589 | 0.642 | 0.697 |
| 1980 | 0.146 | 0.169 | 0.193 | 0.219 | 0.247 | 0.275 | 0.305 | 0.337 | 0.37 | 0.404 | 0.439 | 0.476 | 0.515 | 0.555 |
| 1981 | 0.162 | 0.183 | 0.20 | 0.234 | 0.264 | 0.296 | 0.331 | 0.36 | 0.41 | 0.454 | 0.5 | 0.55 | 0.602 | 0.657 |
| 1982 | 0.112 | 0.171 | 0.225 | 0.275 | 0.321 | 0.362 | 0.399 | 0.432 | 0.461 | 0.485 | 0.505 | 0.52 | 0.531 | 0.538 |
| 1983 | 0.189 | 0.212 | 0.238 | 0.266 | 0.298 | 0.332 | 0.369 | 0.41 | 0.453 | 0.499 | 0.548 | 0.599 | 0.654 | 0.712 |
| 1984 | 0.191 | 0.225 | 0.257 | 0.288 | 0.318 | 0.347 | 0.374 | 0.4 | 0.425 | 0.449 | 0.472 | 0.493 | 0.513 | 0.532 |
| 1985 | 0.144 | 0.18 | 0.231 | 0.272 | 0.31 | 0.346 | 0.38 | 0.412 | 0.441 | 0.469 | 0.494 | 0.517 | 0.538 | 0.557 |
| 1986 | 0.122 | 0.164 | 0.203 | 0.241 | 0.277 | 0.311 | 0.344 | 0.375 | 0.404 | 0.432 | 0.458 | 0.482 | 0.505 | 0.525 |
| 1987 | 0.135 | 0.164 | 0.196 | 0.231 | 0.268 | 0.308 | 0.35 | 0.395 | 0.442 | 0.492 | 0.545 | 0.6 | 0.658 | 0.719 |
| 1988 | 0.111 | 0.14 | 0.183 | 0.218 | 0.252 | 0.286 | 0.319 | 0.352 | 0.384 | 0.415 | 0.446 | 0.476 | 0.505 | 0.534 |
| 1989 | 0.125 | 0.163 | 0.201 | 0.237 | 0.271 | 0.304 | 0.336 | 0.366 | 0.395 | 0.422 | 0.448 | 0.473 | 0.496 | 0.517 |
| 1990 | 0.135 | 0.162 | 0.192 | 0.227 | 0.265 | 0.307 | 0.354 | 0.404 | 0.458 | 0.516 | 0.578 | 0.644 | 0.714 | 0.788 |
| 1991 | 0.133 | 0.172 | 0.208 | 0.241 | 0.272 | 0.3 | 0.326 | 0.349 | 0.369 | 0.386 | 0.401 | 0.413 | 0.423 | 0.43 |
| 1992 | 0.149 | 0.177 | 0.207 | 0.239 | 0.274 | 0.31 | 0.349 | 0.39 | 0.433 | 0.478 | 0.525 | 0.574 | 0.625 | 0.679 |
| 1993 | 0.102 | 0.156 | 0.205 | 0.248 | 0.285 | 0.318 | 0.345 | 0.366 | 0.382 | 0.392 | 0.397 | 0.397 | 0.391 | 0.38 |
| 1994 | 0.175 | 0.198 | 0.227 | 0.261 | 0.301 | 0.346 | 0.397 | 0.453 | 0.515 | 0.582 | 0.654 | 0.732 | 0.816 | 0.905 |
| 1995 | 0.129 | 0.182 | 0.232 | 0.277 | 0.318 | 0.356 | 0.389 | 0.419 | 0.444 | 0.466 | 0.484 | 0.497 | 0.507 | 0.513 |
| 1996 | 0.156 | 0.193 | 0.228 | 0.263 | 0.296 | 0.327 | 0.358 | 0.387 | 0.414 | 0.44 | 0.465 | 0.488 | 0.51 | 0.531 |
| 1997 | 0.154 | 0.197 | 0.237 | 0.275 | 0.311 | 0.345 | 0.376 | 0.406 | 0.433 | 0.458 | 0.481 | 0.501 | 0.519 | 0.536 |
| 1998 | 0.187 | 0.209 | 0.234 | 0.263 | 0.295 | 0.331 | 0.369 | 0.411 | 0.457 | 0.506 | 0.558 | 0.614 | 0.672 | 0.735 |
| 1999 | 0.179 | 0.217 | 0.252 | 0.285 | 0.314 | 0.341 | 0.365 | 0.387 | 0.406 | 0.422 | 0.436 | 0.446 | 0.454 | 0.46 |
| 2000 | 0.14 | 0.189 | 0.25 | 0.311 | 0.368 | 0.428 | 0.384 | 0.456 | 0.613 | 0.533 | 0.412 | 0.517 | 0.631 | 0.784 |
| 2001 | 0.175 | 0.18 | 0.271 | 0.293 | 0.326 | 0.42 | 0.465 | 0.382 | 0.415 | 0.459 | 0.378 | 0.532 | 0.381 | 0.615 |
| 2002 | 0.162 | 0.172 | 0.211 | 0.283 | 0.328 | 0.333 | 0.417 | 0.277 | 0.309 | 0.29 | 0.338 | 0.602 | 0.459 | 0.691 |
| 2003 | 0.16 | 0.187 | 0.247 | 0.294 | 0.342 | 0.326 | 0.35 | 0.594 | 0.505 | 0.576 | 0.23 | 0.48 | 0.632 | 0.455 |
| 2004 | 0.17 | 0.219 | 0.289 | 0.338 | 0.371 | 0.383 | 0.383 | 0.459 | 0.504 | 0.551 | 0.416 | 0.365 | 0.489 | 0.506 |
| 2005 | 0.16 | 0.203 | 0.256 | 0.286 | 0.312 | 0.326 | 0.334 | 0.34 | 0.331 | 0.337 | 0.388 | 0.364 | 0.335 | 0.572 |
| 2006 | 0.179 | 0.194 | 0.224 | 0.297 | 0.293 | 0.318 | 0.302 | 0.315 | 0.337 | 0.39 | 0.391 | 0.768 | 0.395 | 0.517 |
| 2007 | 0.172 | 0.224 | 0.296 | 0.36 | 0.38 | 0.429 | 0.415 | 0.467 | 0.461 | 0.428 | 0.513 | 0.54 | 0.642 | 0.588 |


| year/age $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5 +}$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2008 | 0.148 | 0.189 | 0.248 | 0.279 | 0.291 | 0.386 | 0.397 | 0.373 | 0.368 | 0.35 | 0.451 | 0.34 | 0.588 | 0.54 |
| 2009 | 0.141 | 0.195 | 0.229 | 0.279 | 0.277 | 0.261 | 0.34 | 0.289 | 0.281 | 0.229 | 0.251 | 0.312 | 0.242 | 0.315 |
| 2010 | 0.166 | 0.193 | 0.266 | 0.285 | 0.321 | 0.308 | 0.314 | 0.353 | 0.365 | 0.366 | 0.386 | 0.356 | 0.238 | 0.33 |
| 2011 | 0.215 | 0.213 | 0.276 | 0.362 | 0.413 | 0.368 | 0.383 | 0.318 | 0.521 | 0.287 | 0.46 | 0.297 | 0.368 | 0.292 |
| 2012 | 0.187 | 0.22 | 0.26 | 0.311 | 0.331 | 0.368 | 0.374 | 0.302 | 0.298 | 0.509 | 0.47 | 0.342 | 0.36 | 0.321 |
| 2013 | 0.17 | 0.213 | 0.278 | 0.32 | 0.347 | 0.353 | 0.332 | 0.403 | 0.403 | 0.329 | 0.308 | 0.302 | 0.331 | 0.418 |
| 2014 | 0.17 | 0.196 | 0.269 | 0.328 | 0.369 | 0.397 | 0.402 | 0.452 | 0.444 | 0.49 | 0.416 | 0.524 | 0.607 | 0.427 |
| 2015 | 0.18 | 0.221 | 0.309 | 0.342 | 0.381 | 0.4 | 0.391 | 0.381 | 0.383 | 0.349 | 0.412 | 0.345 | 0.382 | 0.381 |
| 2016 | 0.187 | 0.223 | 0.269 | 0.356 | 0.332 | 0.414 | 0.426 | 0.377 | 0.444 | 0.42 | 0.739 | 0.45 | 0.662 | 0.433 |
| 2017 | 0.177 | 0.239 | 0.323 | 0.386 | 0.495 | 0.493 | 0.503 | 0.472 | 0.44 | 0.588 | 0.331 | 0.373 | 0.457 | 0.465 |
| 2018 | 0.186 | 0.24 | 0.31 | 0.389 | 0.476 | 0.485 | 0.567 | 0.716 | 0.334 | 0.524 | 0.278 | 0.34 | 0.493 | 0.522 |
| 2019 | 0.186 | 0.22 | 0.278 | 0.324 | 0.392 | 0.366 | 0.419 | 0.54 | 0.441 | 0.427 | 0.487 | 0.343 | 0.265 | 0.455 |
| 2020 | 0.169 | 0.219 | 0.273 | 0.335 | 0.378 | 0.371 | 0.347 | 0.358 | 0.448 | 0.308 | 0.452 | 0.422 | 0.564 | 0.339 |
| 2021 | 0.186 | 0.227 | 0.255 | 0.324 | 0.322 | 0.334 | 0.316 | 0.336 | 0.375 | 0.419 | 0.366 | 0.474 | 0.426 | 0.345 |

Table 31.5. Sol.27.7a - Stock weights-at-age (kg).

| year/age |  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 0.13 | 0.153 | 0.178 | 0.204 | 0.232 | 0.26 | 0.29 | 0.321 | 0.353 | 0.387 | 0.422 | 0.458 | 0.495 | 0.533 |
| 1971 | 0.152 | 0.178 | 0.204 | 0.23 | 0.257 | 0.284 | 0.312 | 0.34 | 0.369 | 0.398 | 0.427 | 0.457 | 0.487 | 0.517 |
| 1972 | 0.126 | 0.164 | 0.201 | 0.237 | 0.272 | 0.306 | 0.338 | 0.369 | 0.4 | 0.428 | 0.456 | 0.483 | 0.508 | 0.533 |
| 1973 | 0.151 | 0.178 | 0.204 | 0.23 | 0.256 | 0.283 | 0.309 | 0.335 | 0.361 | 0.387 | 0.413 | 0.439 | 0.464 | 0.49 |
| 1974 | 0.138 | 0.174 | 0.209 | 0.241 | 0.272 | 0.301 | 0.328 | 0.353 | 0.377 | 0.399 | 0.419 | 0.437 | 0.453 | 0.468 |
| 1975 | 0.13 | 0.172 | 0.21 | 0.244 | 0.275 | 0.303 | 0.327 | 0.347 | 0.364 | 0.378 | 0.387 | 0.394 | 0.396 | 0.396 |
| 1976 | 0.12 | 0.161 | 0.2 | 0.239 | 0.276 | 0.313 | 0.348 | 0.383 | 0.416 | 0.449 | 0.48 | 0.511 | 0.541 | 0.569 |
| 1977 | 0.085 | 0.146 | 0.202 | 0.251 | 0.293 | 0.33 | 0.36 | 0.384 | 0.401 | 0.413 | 0.418 | 0.417 | 0.409 | 0.395 |
| 1978 | 0.093 | 0.147 | 0.197 | 0.243 | 0.286 | 0.326 | 0.361 | 0.394 | 0.422 | 0.447 | 0.468 | 0.486 | 0.5 | 0.511 |
| 1979 | 0.134 | 0.165 | 0.199 | 0.234 | 0.271 | 0.311 | 0.352 | 0.395 | 0.441 | 0.488 | 0.537 | 0.589 | 0.642 | 0.697 |
| 1980 | 0.146 | 0.169 | 0.193 | 0.219 | 0.247 | 0.275 | 0.305 | 0.337 | 0.37 | 0.404 | 0.439 | 0.476 | 0.515 | 0.555 |
| 1981 | 0.162 | 0.183 | 0.207 | 0.234 | 0.264 | 0.296 | 0.331 | 0.369 | 0.41 | 0.454 | 0.5 | 0.55 | 0.602 | 0.657 |
| 1982 | 0.112 | 0.171 | 0.225 | 0.275 | 0.321 | 0.362 | 0.399 | 0.432 | 0.461 | 0.485 | 0.505 | 0.52 | 0.531 | 0.538 |
| 1983 | 0.189 | 0.212 | 0.238 | 0.266 | 0.298 | 0.332 | 0.369 | 0.41 | 0.453 | 0.499 | 0.548 | 0.599 | 0.654 | 0.712 |
| 1984 | 0.191 | 0.225 | 0.257 | 0.288 | 0.318 | 0.347 | 0.374 | 0.4 | 0.425 | 0.449 | 0.472 | 0.493 | 0.513 | 0.532 |
| 1985 | 0.144 | 0.189 | 0.231 | 0.272 | 0.31 | 0.346 | 0.38 | 0.412 | 0.441 | 0.469 | 0.494 | 0.517 | 0.538 | 0.557 |
| 1986 | 0.122 | 0.164 | 0.203 | 0.241 | 0.277 | 0.311 | 0.344 | 0.375 | 0.404 | 0.432 | 0.458 | 0.482 | 0.505 | 0.525 |
| 1987 | 0.135 | 0.164 | 0.196 | 0.231 | 0.268 | 0.308 | 0.35 | 0.395 | 0.442 | 0.492 | 0.545 | 0.6 | 0.658 | 0.719 |
| 1988 | 0.111 | 0.147 | 0.183 | 0.218 | 0.252 | 0.286 | 0.319 | 0.352 | 0.384 | 0.415 | 0.446 | 0.476 | 0.505 | 0.534 |
| 1989 | 0.125 | 0.163 | 0.201 | 0.237 | 0.271 | 0.304 | 0.336 | 0.366 | 0.395 | 0.422 | 0.448 | 0.473 | 0.496 | 0.517 |
| 1990 | 0.135 | 0.162 | 0.192 | 0.227 | 0.265 | 0.307 | 0.354 | 0.404 | 0.458 | 0.516 | 0.578 | 0.644 | 0.714 | 0.788 |
| 1991 | 0.133 | 0.172 | 0.208 | 0.241 | 0.272 | 0.3 | 0.326 | 0.349 | 0.369 | 0.386 | 0.401 | 0.413 | 0.423 | 0.43 |
| 1992 | 0.149 | 0.177 | 0.207 | 0.239 | 0.274 | 0.31 | 0.349 | 0.39 | 0.433 | 0.478 | 0.525 | 0.574 | 0.625 | 0.679 |
| 1993 | 0.102 | 0.156 | 0.205 | 0.248 | 0.285 | 0.318 | 0.345 | 0.366 | 0.382 | 0.392 | 0.397 | 0.397 | 0.391 | 0.38 |
| 1994 | 0.175 | 0.198 | 0.227 | 0.261 | 0.301 | 0.346 | 0.397 | 0.453 | 0.515 | 0.582 | 0.654 | 0.732 | 0.816 | 0.905 |
| 1995 | 0.129 | 0.182 | 0.232 | 0.277 | 0.318 | 0.356 | 0.389 | 0.419 | 0.444 | 0.466 | 0.484 | 0.497 | 0.507 | 0.513 |
| 1996 | 0.156 | 0.193 | 0.228 | 0.263 | 0.296 | 0.327 | 0.358 | 0.387 | 0.414 | 0.44 | 0.465 | 0.488 | 0.51 | 0.531 |
| 1997 | 0.154 | 0.197 | 0.237 | 0.275 | 0.311 | 0.345 | 0.376 | 0.406 | 0.433 | 0.458 | 0.481 | 0.501 | 0.519 | 0.536 |
| 1998 | 0.187 | 0.209 | 0.234 | 0.263 | 0.295 | 0.331 | 0.369 | 0.411 | 0.457 | 0.506 | 0.558 | 0.614 | 0.672 | 0.735 |
| 1999 | 0.179 | 0.217 | 0.252 | 0.285 | 0.314 | 0.341 | 0.365 | 0.387 | 0.406 | 0.422 | 0.436 | 0.446 | 0.454 | 0.46 |
| 2000 | 0.124 | 0.158 | 0.23 | 0.303 | 0.345 | 0.41 | 0.385 | 0.478 | 0.707 | 0.633 | 0.362 | 0.602 | 0.571 | 0.784 |
| 2001 | 0.151 | 0.159 | 0.226 | 0.271 | 0.318 | 0.393 | 0.446 | 0.383 | 0.435 | 0.531 | 0.449 | 0.468 | 0.444 | 0.615 |
| 2002 | 0.145 | 0.174 | 0.195 | 0.277 | 0.31 | 0.33 | 0.419 | 0.359 | 0.344 | 0.347 | 0.394 | 0.477 | 0.495 | 0.691 |
| 2003 | 0.144 | 0.174 | 0.207 | 0.249 | 0.311 | 0.327 | 0.341 | 0.498 | 0.374 | 0.422 | 0.258 | 0.403 | 0.617 | 0.455 |
| 2004 | 0.15 | 0.187 | 0.232 | 0.289 | 0.331 | 0.362 | 0.353 | 0.401 | 0.547 | 0.528 | 0.49 | 0.29 | 0.484 | 0.506 |
| 2005 | 0.144 | 0.186 | 0.237 | 0.288 | 0.325 | 0.348 | 0.358 | 0.361 | 0.39 | 0.412 | 0.462 | 0.389 | 0.35 | 0.572 |
| 2006 | 0.152 | 0.177 | 0.213 | 0.276 | 0.289 | 0.315 | 0.314 | 0.324 | 0.339 | 0.36 | 0.363 | 0.546 | 0.38 | 0.518 |
| 2007 | 0.156 | 0.2 | 0.24 | 0.284 | 0.336 | 0.354 | 0.363 | 0.376 | 0.381 | 0.38 | 0.447 | 0.459 | 0.702 | 0.588 |


| year/age $\mathbf{2} \mathbf{3} \mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5 +}$ |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 0.134 | 0.181 | 0.236 | 0.288 | 0.324 | 0.383 | 0.413 | 0.393 | 0.415 | 0.402 | 0.439 | 0.417 | 0.564 | 0.54 |
| 2009 | 0.129 | 0.17 | 0.208 | 0.263 | 0.278 | 0.276 | 0.363 | 0.339 | 0.323 | 0.29 | 0.296 | 0.375 | 0.287 | 0.315 |
| 2010 | 0.158 | 0.165 | 0.228 | 0.256 | 0.3 | 0.292 | 0.286 | 0.347 | 0.325 | 0.321 | 0.297 | 0.299 | 0.272 | 0.331 |
| 2011 | 0.167 | 0.188 | 0.231 | 0.31 | 0.343 | 0.344 | 0.343 | 0.316 | 0.429 | 0.324 | 0.41 | 0.339 | 0.362 | 0.292 |
| 2012 | 0.156 | 0.218 | 0.235 | 0.293 | 0.346 | 0.39 | 0.371 | 0.34 | 0.308 | 0.515 | 0.368 | 0.397 | 0.327 | 0.321 |
| 2013 | 0.149 | 0.2 | 0.248 | 0.288 | 0.329 | 0.342 | 0.35 | 0.388 | 0.349 | 0.313 | 0.396 | 0.377 | 0.337 | 0.418 |
| 2014 | 0.111 | 0.183 | 0.24 | 0.302 | 0.343 | 0.371 | 0.377 | 0.388 | 0.423 | 0.444 | 0.37 | 0.402 | 0.428 | 0.427 |
| 2015 | 0.153 | 0.194 | 0.246 | 0.303 | 0.353 | 0.384 | 0.394 | 0.392 | 0.416 | 0.394 | 0.449 | 0.379 | 0.447 | 0.381 |
| 2016 | 0.127 | 0.2 | 0.244 | 0.332 | 0.337 | 0.397 | 0.413 | 0.384 | 0.412 | 0.401 | 0.508 | 0.431 | 0.478 | 0.433 |
| 2017 | 0.152 | 0.212 | 0.268 | 0.322 | 0.42 | 0.405 | 0.456 | 0.448 | 0.407 | 0.511 | 0.373 | 0.525 | 0.454 | 0.465 |
| 2018 | 0.149 | 0.206 | 0.273 | 0.354 | 0.428 | 0.49 | 0.529 | 0.6 | 0.397 | 0.481 | 0.405 | 0.335 | 0.429 | 0.522 |
| 2019 | 0.155 | 0.202 | 0.259 | 0.317 | 0.39 | 0.417 | 0.451 | 0.553 | 0.562 | 0.378 | 0.505 | 0.309 | 0.3 | 0.455 |
| 2020 | 0.119 | 0.202 | 0.245 | 0.305 | 0.35 | 0.381 | 0.356 | 0.387 | 0.492 | 0.369 | 0.439 | 0.453 | 0.44 | 0.339 |
| 2021 | 0.127 | 0.196 | 0.236 | 0.297 | 0.329 | 0.356 | 0.343 | 0.341 | 0.367 | 0.433 | 0.336 | 0.463 | 0.424 | 0.345 |

Table 31.6a. Sol.27.7a - Discard rates for the main fleets operational in the Irish Sea (Belgian, UK and Irish beam trawl, UK and Irish otter trawl, UK and Irish Nephrops trawl).

|  | BEL | UK |  |  |  |  |  | IRL |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gear | TBB | TBB | Отв | TWIN OTB | NEPH OTB | TWIN NEPH | Other | TBB | NEPH OTB | OTB DEF |
| Landings ( t ) | 716 | 284 | 61 | 4 | 25 | 6 | Na | 427 | / | / |
| Discard ratio | 0.05 | 0.08 | 0.05 | 0.01 | 0.08 | 0.02 | Na | 0.02 | / | / |
| years | 2007-2009 | $\begin{gathered} 2002 \\ 2005-2007 \end{gathered}$ | 2002-2009 | 2003,2004,2007 | $\begin{gathered} 2003, \\ 2006-2009 \end{gathered}$ | 2002,2003,2008 | Na | 2003-2009 | / | / |
| Landings (t) 2010 | 210.917 | 1.721 | 1.071 | 0.014 | 3.329 | 0.501 | 0.741 | 38.283 | 5.327 | 3.632 |
| Discard ratio 2010 | 0.04 | Na | 0.00 | Na | 0.05 | Na | Na | 0.05 | 0.16* | 0.39* |
| Landings (t) 2011 | 239.483 | 13.662 | 2.866 | 0.05 | 5.201 | 0.414 | 0.821 | 32.514 | 10.116 | 5.581 |
| Discard ratio 2011 | 0.04 | Na | 0.02 | Na | 0.00 | Na | Na | 0.003 | 0.16* | 0.00 |

* It should be noted that the $\mathbf{1 6 \%}$ discard rate for 2010-2011 of the Irish Nephrops fleet and the $\mathbf{3 9 \%}$ discard rate for 2010 of the Irish otter trawl fleet only accounts for respectively $\mathbf{1 . 9} \%$, $\mathbf{3} .1 \%$ and $1.3 \%$ of the total international landings.

Table 31.6b. Sol.27.7a - Discard rates.


total L L corresponding\% coverage of L total D rate with discard info

| Country Year BE | Landings (L) (t) |  |  | Discards (D) (t) |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | TBB | OTB | other |  |  |
| 2012 | 286.44 | 227.01 | 0.79 | 17.22 | 0.071 |
| 2013 | 142.74 | 107.35 | 0.75 | 8.54 | 0.074 |
| 2014 | 91.93 | 45.74 | 0.50 | 4.09 | 0.082 |
| 2015 | 71.88 | 42.89 | 0.60 | 3.74 | 0.080 |
| 2016 | 32.61 | 21.08 | 0.65 | 0.64 | 0.029 |
| 2017 | 30.38 | 18.99 | 0.63 | 0.64 | 0.033 |
| 2018 | 34.02 | 22.50 | 0.66 | 0.97 | 0.041 |
| 2019 | 392.31 | 351.42 | 0.90 | 55.31 | 0.136 |
| 2020 | 396.45 | 293.32 | 0.74 | 41.00 | 0.123 |
| 2021 | 609.58 | 455.64 | 0.75 | 44.35 | 0.089 |
| average 19-21 |  |  |  |  | 0.116 |

Table 31.7a. Sol.27.7a - Effort series.

|  | Belgium <br> beam ${ }^{1}$ | UK(E\&W) <br> beam ${ }^{2}$ | beam ${ }^{3}$ | otter ${ }^{2}$ | otter ${ }^{3}$ | Ireland otter ${ }^{4}$ | beam ${ }^{4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Whole | Whole | Whole | Whole | Whole | Whole | Whole |
|  | year | year | year | year | year | Year | Year |
| 1972 | - | - | - | 128.4 | - | - | - |
| 1973 | - | - | - | 147.6 | - | - | - |
| 1974 | - | - | - | 115.2 | - | - | - |
| 1975 | 28.4 | - | - | 130.7 | - | - | - |
| 1976 | 24.9 | - | - | 122.3 | - | - | - |
| 1977 | 22.1 | - | - | 101.9 | - | - | - |
| 1978 | 17.5 | 0.9 | - | 89.1 | - | - | - |
| 1979 | 20.4 | 1.7 | - | 89.9 | - | - | - |
| 1980 | 32.0 | 4.3 | - | 107.0 | - | - | - |
| 1981 | 36.5 | 6.4 | - | 107.1 | - | - | - |
| 1982 | 26.5 | 5.5 | - | 127.2 | - | - | - |
| 1983 | 28.7 | 2.8 | 0.0 | 88.1 | 1716.5 | - | - |
| 1984 | 17.5 | 4.1 | 263.0 | 103.1 | 7932.1 | - | - |
| 1985 | 27.0 | 7.4 | 428.1 | 102.9 | 6930.8 | - | - |
| 1986 | 44.5 | 17.0 | 1122.9 | 90.3 | 6693.2 | - | - |
| 1987 | 51.6 | 22.0 | 1178.5 | 130.6 | 9008.9 | - | - |
| 1988 | 38.2 | 18.6 | 1019.2 | 132.0 | 8292.4 | - | - |
| 1989 | 42.2 | 25.3 | 1344.5 | 139.5 | 16161.4 | - | - |
| 1990 | 42.4 | 31.0 | 1473.1 | 117.1 | 7724.5 | - | - |
| 1991 | 17.1 | 25.8 | 1211.3 | 107.3 | 7081.1 | - | - |
| 1992 | 25.1 | 23.4 | 908.1 | 96.8 | 6671.8 | - | - |
| 1993 | 23.9 | 21.5 | 826.9 | 78.9 | 6013.1 | - | - |
| 1994 | 32.5 | 20.1 | 1451.6 | 43.0 | 3060.0 | - | - |
| 1995 | 28.6 | 20.9 | 1429.4 | 43.1 | 3357.0 | 80.3 | 8.6 |
| 1996 | 23.2 | 13.3 | 894.3 | 42.2 | 3085.1 | 64.8 | 6.3 |
| 1997 | 30.7 | 10.8 | 784.4 | 39.9 | 2903.3 | 92.2 | 9.9 |
| 1998 | 24.7 | 10.4 | 696.0 | 36.9 | 2620.6 | 93.5 | 11.6 |
| 1999 | 22.7 | 11.0 | 778.9 | 22.9 | 1803.5 | 110.3 | 14.7 |
| 2000 | 26.0 | 6.3 | 410.7 | 27.0 | 2034.9 | 82.7 | 11.4 |
| 2001 | 36.8 | 12.5 | 767.4 | 32.8 | 2352.9 | 77.5 | 13.1 |
| 2002 | 47.0 | 8.0 | 535.1 | 24.8 | 1774.0 | 77.9 | 17.7 |
| 2003 | 44.3 | 14.0 | 863.7 | 23.9 | 1728.3 | 73.9 | 18.7 |
| 2004 | 32.3 | 7.4 | 419.9 | 23.5 | 1727.0 | 72.5 | 14.2 |
| 2005 | 37.5 | 11.4 | 627.8 | 16.7 | 1313.6 | 68.3 | 14.7 |
| 2006 | 24.8 | 4.6 | 280.1 | 5.2 | 478.5 | 66.2 | 12.2 |
| 2007 | 19.5 | 3.2 | 193.5 | 4.4 | 397.2 | 74.1 | 14.2 |


|  | Belgium <br> beam ${ }^{1}$ | UK(E\&W) <br> beam ${ }^{2}$ | beam ${ }^{3}$ | otter ${ }^{2}$ | otter ${ }^{3}$ | Ireland otter ${ }^{4}$ | beam ${ }^{4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 10.3 | 1.3 | 98.0 | 2.7 | 320.4 | 58.8 | 9.5 |
| 2009 | 11.7 | 0.5 | 24.9 | 1.5 | 157.7 | 42.8 | 7.6 |
| 2010 | 11.3 | 0.2 | 10.2 | 1.4 | 151.0 | 45.8 | 9.4 |
| 2011 | 12.4 | 1.6 | 91.2 | 0.7 | 72.7 | 54.5 | 8.1 |
| 2012 | 10.9 | 0.9 | 60.7 | 0.4 | 85.0 | 58.3 | 7.2 |
| 2013 | 7.0 | 0.0 | 1.3 | 0.3 | 31.9 | 42.6 | 5.0 |
| 2014 | 3.9 | - | 0.4 | - | 16.1 | 47.7 | 6.0 |
| 2015 | 3.5 | - | 0.9 | - | 0.0 | 39.8 | 8.3 |
| 2016 | 1.8 | - | 3.9 | - | 0.0 | 33.4 | 7.9 |
| 2017 | 3.0 | - | 0.0 | - | 160.7 | 12.1 | 7.5 |
| 2018 | 2.5 | - | 0.0 | - | 238.1 | 13.6 | 9.6 |
| 2019 | 10.1 | - | 7.0 | - | 247.2 | 17.2 | 13.3 |
| 2020 | 9.5 | - | 199.0 | - | 72.6 | 13.3 | 10.4 |
| 2021 | 9.9 | - | 222.6 | - | 193.6 | 13.2 | 8.9 |

${ }^{1000}$ hours fishing.
${ }^{2} 000$ 'hours fished (GRT corrected $>40$ vessels).
${ }^{3}$ days fished.
${ }^{4} 000$ 'hours.

* Provisional.

Table 31.7b. Sol.27.7a - LPUE.

|  | Belgium <br> beam ${ }^{1}$ | UK(E\&W) |  |  | UK |  |  | Ireland |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | beam ${ }^{3}$ | beam ${ }^{2}$ | otter ${ }^{3}$ | otter ${ }^{2}$ | beam surv |  | otter ${ }^{1}$ | beam ${ }^{1}$ |
|  | Whole | Whole | Whole | Whole | Whole | Sept | March | Whole | Whole |
|  | year | year | year | year | year |  |  | year | year |
| 1972 | - | - | - | 1.06 | - | - | - | - | - |
| 1973 | - | - | - | 1.06 | - | - | - | - | - |
| 1974 | - | - | - | 1.09 | - | - | - | - | - |
| 1975 | 21.39 | - | - | 1.39 | - | - | - | - | - |
| 1976 | 23.13 | - | - | 0.94 | - | - | - | - | - |
| 1977 | 19.79 | - | - | 0.80 | - | - | - | - | - |
| 1978 | 18.10 | 34.32 | - | 1.04 | - | - | - | - | - |
| 1979 | 33.41 | 32.01 | - | 1.43 | - | - | - | - | - |
| 1980 | 28.18 | 31.70 | - | 1.01 | - | - | - | - | - |
| 1981 | 22.16 | 21.32 | - | 0.75 | - | - | - | - | - |
| 1982 | 22.01 | 29.94 | - | 0.53 | - | - | - | - | - |
| 1983 | 13.88 | 37.31 | 0.0 | 0.57 | 150.2 | - | - | - | - |
| 1984 | 22.47 | 16.24 | 2851.4 | 0.71 | 119.3 | - | - | - | - |
| 1985 | 20.58 | 17.34 | 2956.3 | 0.56 | 135.7 | - | - | - | - |
| 1986 | 19.12 | 19.23 | 3925.7 | 0.84 | 174.9 | - | - | - | - |
| 1987 | 17.73 | 14.82 | 3726.9 | 0.77 | 144.9 | - | - | - | - |
| 1988 | 21.29 | 11.81 | 2673.3 | 0.46 | 80.3 | 161.92 | - | - | - |
| 1989 | 21.93 | 9.17 | 1750.6 | 0.70 | 138.9 | 150.07 | - | - | - |
| 1990 | 17.52 | 9.52 | 2300.9 | 0.61 | 119.7 | 196.90 | - | - | - |
| 1991 | 18.70 | 10.43 | 2420.9 | 1.12 | 177.4 | 175.76 | - | - | - |
| 1992 | 19.21 | 9.50 | 2763.0 | 1.02 | 126.0 | 162.64 | - | - | - |
| 1993 | 19.97 | 7.60 | 1879.8 | 0.54 | 69.1 | 100.16 | 104.7 | - | - |
| 1994 | 19.06 | 11.76 | 1479.9 | 0.74 | 88.1 | 110.71 | 91.9 |  | - |
| 1995 | 18.12 | 14.96 | 1721.1 | 0.95 | 142.3 | 92.04 | 79.3 | 0.38 | 12.69 |
| 1996 | 17.72 | 9.44 | 1471.7 | 0.53 | 47.7 | 89.48 | - | 0.25 | 14.94 |
| 1997 | 16.62 | 10.49 | 961.8 | 0.73 | 103.2 | 155.79 | 63.3 | 0.23 | 8.53 |
| 1998 | 18.96 | 8.42 | 907.8 | 0.48 | 50.5 | 144.97 | 89.3 | 0.38 | 7.77 |
| 1999 | 19.47 | 9.94 | 1124.9 | 0.60 | 64.8 | 116.02 | - | 0.29 | 9.22 |
| 2000 | 15.52 | 12.90 | 1604.7 | 0.44 | 34.6 | 130.70 | - | 0.29 | 8.49 |
| 2001 | 15.02 | 11.72 | 1537.4 | 0.15 | 23.4 | 96.87 | - | 0.38 | 7.86 |
| 2002 | 14.95 | 16.73 | 1484.3 | 1.48 | 98.8 | 76.73 | - | 0.32 | 4.67 |
| 2003 | 15.41 | 13.20 | 1351.6 | 0.15 | 340.4 | 88.55 | - | 0.34 | 4.20 |
| 2004 | 16.25 | 13.86 | 941.7 | 0.17 | 27.6 | 98.92 | - | 0.14 | 4.31 |
| 2005 | 17.52 | 9.14 | 1199.9 | 0.19 | 21.3 | 48.91 | - | 0.16 | 4.70 |


|  | Belgium <br> beam ${ }^{1}$ | UK(E\&W) <br> beam ${ }^{3}$ | beam ${ }^{2}$ | otter ${ }^{3}$ | otter ${ }^{2}$ | UK <br> beam survey ${ }^{4}$ |  | Ireland <br> otter ${ }^{1}$ | beam ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 16.32 | 7.83 | 826.1 | 0.52 | 34.8 | 52.63 | - | 0.16 | 6.00 |
| 2007 | 14.32 | 16.38 | 1629.9 | 0.42 | 21.4 | 53.05 | - | 0.37 | 6.37 |
| 2008 | 19.85 | 15.25 | 887.4 | 0.30 | 16.4 | 50.67 | - | 0.20 | 6.08 |
| 2009 | 19.96 | 18.88 | 1201.2 | 0.22 | 13.6 | 45.75 | - | 0.28 | 4.53 |
| 2010 | 18.68 | 13.90 | 262.3 | 0.46 | 17.8 | 27.80 | - | 0.19 | 4.09 |
| 2011 | 19.34 | 4.45 | 322.5 | 0.18 | 13.7 | 36.97 | - | 0.30 | 4.13 |
| 2012 | 19.61 | 4.27 | 99.9 | 0.08 | 4.4 | 26.47 | - | 0.14 | 5.41 |
| 2013 | 13.23 | - | 27.7 | 0.10 | 16.3 | 31.65 | - | 0.22 | 6.27 |
| 2014 | 9.16 | - | 0.0 | - | 13.0 | 41.14 | - | 0.14 | 5.40 |
| 2015 | 9.24 | - | 146.1 | - | 34.2 | 58.88 | - | 0.18 | 3.14 |
| 2016 | 6.81 | - | 0.0 | - | 21.3 | 69.35 | - | 0.18 | 1.17 |
| 2017 | 3.81 | - | 0.0 | - | 5.6 | 64.24 | - | 0.36 | 1.23 |
| 2018 | 5.36 | - | 0.0 | - | 12.6 | 78.51 | - | 0.28 | 1.49 |
| 2019 | 32.26 | - | 124.8 | - | 48.5 | 118.66 | - | 0.63 | 2.23 |
| 2020 | 30.81 | - | 1305.4 | - | 97.2 | - | - | 0.67 | 2.92 |
| 2021* | 46.17 | - | 1079.7 | - | 192.7 | 63.60 | - | 0.72 | 5.79 |

${ }^{\text {all }}$ LPUE values in $\mathrm{Kg} / \mathrm{hr}$.
${ }^{1} \mathrm{Kg} / \mathrm{hr}$.
${ }^{2} \mathrm{Kg} /$ day.
${ }^{3} \mathrm{Kg} / 000$ 'hr fished (GRT corrected $>40$ ' vessels).
${ }^{4} \mathrm{Kg} / 100 \mathrm{~km}$ fished.

Table 31.8. Sol.27.7a - Tuning series (values in bold are used in the assessment).

| BE-CBT$1975$ | Belgium Commercial Beam trawl (Effort = Corrected formula) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2005 |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |
| 4 | 14 |  |  |  |  |  |  |  |  |  |  |
| 12.3 | 1045 | 275 | 393 | 69 | 105 | 94 | 61 | 72 | 11 | 15 | 64 |
| 11.8 | 568 | 1066 | 80 | 263 | 64 | 58 | 35 | 5 | 56 | 5 | 5 |
| 10.7 | 434 | 307 | 509 | 76 | 93 | 45 | 23 | 20 | 2 | 35 | 32 |
| 9.9 | 169 | 304 | 155 | 258 | 41 | 90 | 12 | 29 | 12 | 7 | 17 |
| 11.2 | 1455 | 510 | 323 | 193 | 162 | 37 | 36 | 9 | 41 | 0 | 0 |
| 16.7 | 958 | 1644 | 296 | 268 | 247 | 210 | 30 | 64 | 31 | 14 | 7 |
| 22.6 | 909 | 721 | 998 | 62 | 92 | 44 | 161 | 13 | 92 | 10 | 8 |
| 19.5 | 451 | 608 | 378 | 394 | 52 | 64 | 11 | 29 | 24 | 5 | 0 |
| 20.5 | 259 | 310 | 394 | 238 | 216 | 44 | 38 | 28 | 49 | 3 | 26 |
| 12 | 107 | 204 | 143 | 188 | 91 | 121 | 2 | 1 | 4 | 14 | 0 |
| 19.6 | 606 | 171 | 186 | 99 | 150 | 125 | 83 | 27 | 13 | 4 | 23 |
| 38 | 1531 | 468 | 138 | 135 | 90 | 104 | 69 | 69 | 20 | 8 | 21 |
| 43.2 | 1527 | 881 | 297 | 167 | 69 | 39 | 54 | 59 | 40 | 13 | 9 |
| 30.5 | 2027 | 1012 | 480 | 21 | 33 | 37 | 34 | 42 | 35 | 0 | 7 |
| 34 | 376 | 2423 | 751 | 250 | 59 | 15 | 9 | 2 | 14 | 0 | 1 |
| 36.1 | 307 | 223 | 1263 | 276 | 142 | 13 | 9 | 11 | 11 | 8 | 5 |
| 13.8 | 253 | 78 | 60 | 588 | 115 | 40 | 16 | 1 | 1 | 11 | 3 |
| 23.9 | 298 | 330 | 68 | 40 | 203 | 93 | 36 | 12 | 0 | 0 | 0 |
| 24.5 | 862 | 253 | 149 | 89 | 79 | 160 | 66 | 77 | 0 | 0 | 0 |
| 31 | 680 | 786 | 164 | 103 | 39 | 117 | 58 | 19 | 15 | 0 | 7 |
| 26.2 | 729 | 366 | 410 | 52 | 27 | 6 | 28 | 15 | 6 | 11 | 3 |
| 21.6 | 537 | 334 | 241 | 219 | 53 | 13 | 11 | 14 | 9 | 7 | 2 |
| 28.5 | 270 | 376 | 180 | 162 | 134 | 28 | 27 | 15 | 9 | 8 | 1 |
| 23.3 | 248 | 146 | 142 | 89 | 73 | 62 | 20 | 20 | 9 | 10 | 3 |
| 21.7 | 693 | 199 | 65 | 50 | 37 | 21 | 17 | 9 | 6 | 4 | 6 |
| 18.6 | 685 | 220 | 107 | 31 | 15 | 33 | 13 | 7 | 9 | 0.6 | 8 |


| 30.5 | 600 | 284 | 248 | 39 | 35 | 44 | 33 | 1 | 3 | 0.2 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 38.6 | 1138 | 814 | 349 | 109 | 30 | 9 | 2 | 1 | 1 | 1 | 0 |
| 24.45 | 724 | 436 | 196 | 84 | 20 | 7 | 2 | 1 | 0 | 2 | 1 |
| 25.58 | 313 | 197 | 159 | 47 | 12 | 11 | 6 | 3 | 0 | 0 | 0 |
| 32.15 | 505 | 342 | 156 | 71 | 87 | 9 | 7 | 1 | 13 | 2 | 1 |
| UK(E\&W)-BTS-Q3 |  | September beam trawl survey |  |  |  |  |  |  |  |  |  |
| 1988 | 2021 |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.75 | 0.85 |  |  |  |  |  |  |  |  |
| 1 | 9 |  |  |  |  |  |  |  |  |  |  |
| 100.062 | 118 | 196 | 180 | 410 | 76 | 40 | 4 | 0 | 4 |  |  |
| 129.71 | 218 | 304 | 180 | 74 | 284 | 56 | 32 | 8 | 6 |  |  |
| 128.969 | 1712 | 534 | 122 | 42 | 88 | 194 | 40 | 20 | 6 |  |  |
| 123.78 | 148 | 1286 | 122 | 26 | 16 | 14 | 55 | 19 | 7 |  |  |
| 129.525 | 220 | 309 | 657 | 142 | 34 | 22 | 7 | 75 | 17 |  |  |
| 131.192 | 83 | 330 | 143 | 211 | 40 | 17 | 7 | 16 | 36 |  |  |
| 124.892 | 60 | 408 | 203 | 73 | 132 | 49 | 11 | 13 | 6 |  |  |
| 126.004 | 246 | 154 | 253 | 110 | 30 | 67 | 12 | 5 | 5 |  |  |
| 126.004 | 886 | 126 | 32 | 76 | 46 | 23 | 31 | 8 | 2 |  |  |
| 126.004 | 1158 | 577 | 72 | 24 | 55 | 27 | 16 | 30 | 7 |  |  |
| 126.004 | 539 | 716 | 292 | 18 | 6 | 24 | 23 | 5 | 18 |  |  |
| 126.004 | 385 | 293 | 255 | 203 | 29 | 8 | 26 | 5 | 6 |  |  |
| 126.004 | 354 | 464 | 147 | 219 | 91 | 13 | 2 | 13 | 6 |  |  |
| 126.004 | 91 | 284 | 192 | 65 | 96 | 63 | 6 | 3 | 12 |  |  |
| 126.004 | 205 | 61 | 121 | 126 | 42 | 79 | 49 | 2 | 1 |  |  |
| 126.004 | 242 | 210 | 51 | 97 | 81 | 40 | 43 | 26 | 1 |  |  |
| 126.004 | 406 | 240 | 119 | 27 | 77 | 45 | 41 | 17 | 19 |  |  |
| 122.298 | 53 | 165 | 69 | 25 | 13 | 35 | 25 | 4 | 6 |  |  |
| 126.004 | 107 | 110 | 90 | 45 | 36 | 9 | 16 | 15 | 10 |  |  |
| 126.004 | 125 | 93 | 49 | 57 | 41 | 11 | 4 | 6 | 12 |  |  |
| 122.298 | 126 | 126 | 60 | 21 | 43 | 23 | 6 | 2 | 9 |  |  |


| 126.004 | 60 | 150 | 68 | 40 |  | 19 | 30 | 12 | 7 | 1 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 126.004 | 26 | 60 | 74 | 37 |  | 17 | 5 | 9 | 9 | 3 |  |  |  |
| 122.298 | 88 | 35 | 62 | 68 |  | 35 | 12 | 4 | 13 | 6 |  |  |  |
| 122.298 | 22 | 49 | 16 | 46 |  | 25 | 12 | 11 | 2 | 6 |  |  |  |
| 126.004 | 75 | 57 | 36 | 21 |  | 33 | 18 | 21 | 9 | 1 |  |  |  |
| 126.004 | 172 | 43 | 22 | 35 |  | 14 | 26 | 21 | 14 | 6 |  |  |  |
| 126.004 | 421 | 150 | 41 | 20 |  | 23 | 5 | 15 | 29 | 8 |  |  |  |
| 122.298 | 129 | 363 | 91 | 29 |  | 20 | 24 | 8 | 8 | 9 |  |  |  |
| 126.004 | 237 | 101 | 177 | 56 |  | 24 | 15 | 9 | 7 | 7 |  |  |  |
| 126.004 | 268 | 275 | 75 | 144 |  | 38 | 21 | 9 | 6 | 9 |  |  |  |
| 126.004 | 1018 | 224 | 167 | 68 |  | 132 | 37 | 11 | 4 | 3 |  |  |  |
| - | - | - | - | - |  | - | - | - | - | - |  |  |  |
| 126.004 | 93 | 93 | 282 | 65 |  | 38 | 22 | 50 | 2 | 0 |  |  |  |
| UK(E\&W)-BTS-Q1 |  | March beam trawl survey |  |  |  |  |  |  |  |  |  |  |  |
| 1993 | 1999 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.15 | 0.25 |  |  |  |  |  |  |  |  |  |  |
| 1 | 9 |  |  |  |  |  |  |  |  |  |  |  |  |
| 126.931 | 18 | 337 | 147 | 332 | 73 | 15 | 17 | 10 | 41 |  |  |  |  |
| 115.442 | 8 | 354 | 208 | 69 | 151 | 51 | 14 | 11 | 9 |  |  |  |  |
| 126.189 | 24 | 96 | 186 | 140 | 30 | 104 | 27 | 10 | 8 |  |  |  |  |
| 134.343 | 651 | 114 | 49 | 110 | 78 | 32 | 54 | 10 | 12 |  |  |  |  |
| 121.742 | 130 | 417 | 33 | 17 | 69 | 23 | 11 | 46 | 17 |  |  |  |  |
| 130.081 | 47 | 421 | 330 | 39 | 19 | 48 | 27 | 12 | 37 |  |  |  |  |
| 130.822 | 45 | 227 | 284 | 177 | 14 | 4 | 34 | 12 | 7 |  |  |  |  |
| UK(E\&W)-CBT |  | UK Commercial Beam trawl |  |  |  |  |  |  |  |  |  |  |  |
| 1991 | 2013 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2 | 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 25.838 | 267 | 426 | 212 | 84 | 58 | 218 | 53 | 34 | 4 | 1 | 2 | 1 | 0 |


| 23.399 | 36 | 460 | 176 | 68 | 37 | 32 | 121 | 34 | 38 | 3 | 1 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 21.503 | 11 | 74 | 355 | 98 | 36 | 48 | 25 | 34 | 13 | 22 | 5 | 2 | 4 |
| 20.145 | 24 | 228 | 150 | 234 | 87 | 17 | 25 | 19 | 42 | 10 | 17 | 1 | 0 |
| 20.392 | 47 | 239 | 231 | 130 | 199 | 55 | 11 | 22 | 5 | 34 | 10 | 11 | 3 |
| 13.32 | 0 | 13 | 109 | 98 | 49 | 100 | 37 | 9 | 8 | 6 | 14 | 8 | 3 |
| 10.76 | 0 | 111 | 50 | 81 | 58 | 24 | 46 | 34 | 12 | 12 | 0 | 8 | 1 |
| 10.386 | 43 | 219 | 40 | 28 | 49 | 31 | 12 | 22 | 11 | 9 | 2 | 1 | 0 |
| 11.016 | 53 | 115 | 134 | 12 | 15 | 25 | 10 | 9 | 14 | 9 | 0 | 1 | 2 |
| 6.275 | 16 | 90 | 84 | 82 | 9 | 6 | 10 | 5 | 5 | 7 | 2 | 1 | 1 |
| 12.495 | 33 | 184 | 100 | 145 | 107 | 12 | 4 | 17 | 12 | 10 | 6 | 4 | 2 |
| 8.017 | 4 | 63 | 152 | 50 | 79 | 47 | 5 | 4 | 6 | 3 | 1 | 1 | 1 |
| 13.996 | 28 | 63 | 178 | 149 | 78 | 52 | 72 | 7 | 5 | 8 | 3 | 7 | 14 |
| 7.396 | 54 | 61 | 29 | 43 | 25 | 12 | 10 | 5 | 1 | 1 | 4 | 0 | 1 |
| 11.406 | 10 | 81 | 44 | 16 | 45 | 37 | 17 | 10 | 17 | 3 | 0 | 3 | 3 |
| 4.649 | 7 | 28 | 33 | 11 | 5 | 10 | 12 | 7 | 9 | 5 | 2 | 0 | 1 |
| 3.197 | 22 | 20 | 34 | 17 | 6 | 1 | 7 | 7 | 6 | 3 | 2 | 1 | 1 |
| 1.302 | 1 | 11 | 5 | 7 | 12 | 1 | 2 | 4 | 3 | 4 | 0 | 3 | 1 |
| 0.462 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.186 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1.564 | 0 | 3 | 6 | 3 | 3 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 0.849 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK(E\&W)-COT |  | UK Commercial Otter trawl |  |  |  |  |  |  |  |  |  |  |  |
| 1991 | 2013 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2 | 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 107.3 | 265 | 155 | 63 | 29 | 19 | 71 | 20 | 11 | 2 | 0 | 1 | 1 | 1 |
| 96.8 | 16 | 224 | 69 | 22 | 16 | 10 | 36 | 10 | 10 | 1 | 0 | 0 | 0 |
| 78.9 | 9 | 27 | 77 | 19 | 3 | 7 | 4 | 5 | 1 | 2 | 0 | 0 | 0 |


| 43 | 4 | 66 | 34 | 50 | 20 | 3 | 4 | 4 | 7 | 1 | 2 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 43.1 | 17 | 50 | 34 | 15 | 24 | 7 | 1 | 2 | 0 | 2 | 1 | 1 | 0 |
| 42.2 | 2 | 5 | 18 | 12 | 7 | 12 | 4 | 1 | 1 | 1 | 1 | 1 | 1 |
| 39.9 | 14 | 15 | 7 | 14 | 9 | 3 | 7 | 3 | 1 | 1 | 0 | 1 | 0 |
| 36.9 | 5 | 24 | 5 | 3 | 5 | 3 | 2 | 2 | 1 | 1 | 0 | 0 | 0 |
| 22.8 | 5 | 15 | 12 | 2 | 0 | 2 | 1 | 1 | 1 | 1 | 0 | 0 | 0 |
| 27 | 2 | 12 | 9 | 8 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 32.9 | 3 | 10 | 6 | 8 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 24.8 | 0 | 8 | 16 | 3 | 5 | 3 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
| 23.9 | 1 | 2 | 6 | 4 | 2 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 23.5 | 3 | 5 | 3 | 4 | 3 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 16.7 | 2 | 4 | 2 | 1 | 2 | 2 | 1 | 1 | 1 | 0 | 0 | 0 | 0 |
| 5.2 | 1 | 2 | 4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 0 | 0 | 0 |
| 4.4 | 1 | 1 | 2 | 2 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 |
| 2.7 | 0 | 1 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1.54 | 0 | 0 | 0.2 | 0.3 | 0.1 | 0.2 | 0.2 | 0 | 0 | 0.1 | 0 | 0 | 0 |
| 1.42 | 0 | 0.1 | 0.2 | 0.3 | 0.1 | 0.1 | 0.2 | 0.1 | 0 | 0.1 | 0.1 | 0.1 | 0 |
| 0.686 | 0 | 0.1 | 0.1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.241 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.272 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| IR-COT | Irish Commercial Otter trawl |  |  |  |  |  |  |  |  |  |  |  |  |
| 1995 | 2005 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2 | 10 |  |  |  |  |  |  |  |  |  |  |  |  |
| 70682 | 6.8 | 17.7 | 25.5 | 9.2 | 25.8 | 3.6 | 0.8 | 1.5 | 1.9 | 1995 |  |  |  |
| 58166 | 0 | 5.7 | 12.9 | 12.7 | 4.7 | 4.7 | 2.2 | 0.2 | 0 | 1996 |  |  |  |
| 75029 | 27.8 | 10.2 | 4.1 | 9.2 | 6.4 | 3.5 | 3.9 | 1 | 0.2 | 1997 |  |  |  |
| 81073 | 5.5 | 40.7 | 14.7 | 6.6 | 12.3 | 5.4 | 2.7 | 4.1 | 1 | 1998 |  |  |  |
| 93221 | 26.6 | 36.8 | 30.9 | 5.1 | 3.8 | 5.3 | 2.4 | 0.5 | 1.2 | 1999 |  |  |  |
| 64320 | 1.6 | 13.2 | 13.4 | 11 | 3.4 | 1.1 | 1 | 0.4 | 0 | 2000 |  |  |  |


| 77541 | 0.2 | 6.1 | 18.6 | 18.6 | 10.8 | 2.1 | 4.1 | 1.3 | 0.3 | 2001 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 39996 | 20.3 | 20 | 30.2 | 16.4 | 8.2 | 2.9 | 2.4 | 1.4 | 0.5 | 2002 |
| 73854 | 0.9 | 35.9 | 21.7 | 9.8 | 3.3 | 0.5 | 0.8 | 0.2 | 0.2 | 2003 |
| 72507 | 9 | 15.1 | 4.1 | 3.2 | 1.9 | 1.6 | 0.3 | 0.2 | 0.1 | 2004 |
| \#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\# |  |  |  |  |  |  |  |  |  |  |
| 31142 | 4 | 1.7 | 1.6 | 1.6 | 0.6 | 0.1 | 0 | 0 | 0 | 2005 |

\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#

Please note the 2005 data are based only on Q3 and Q4 data and have not been raised to annual effort.
It should not be included as part of this time-series.

## Table 31.9. Sol.27.7a - Diagnostics.

FLR XSA Diagnostics 2022-04-22 10:58:04
CPUE data from indices
Catch data for 52 years. 1970 to 2021. Ages 2 to 8 .
fleet first age last age first year last year alpha beta
$\begin{array}{lllllll}1 \text { UK (E\&W)-BTS-Q3 } & 2 & 7 & 1988 & 2021 & 0.75 & 0.85\end{array}$

Time series weights :

Tapered time weighting not applied
Catchability analysis :

Catchability independent of size for all ages
Catchability independent of age for ages > 4
Terminal population estimation :
Survivor estimates shrunk towards the mean $F$
of the final 5 years or the 3 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1.5$

Minimum standard error for population
estimates derived from each fleet $=0.3$
prior weighting not applied
Regression weights year
age 2012201320142015201620172018201920202021

Fishing mortalities
age $\begin{array}{llllllllll}2012 & 2013 & 2014 & 2015 & 2016 & 2017 & 2018 & 2019 & 2020 & 2021\end{array}$ $20.0210 .0390 .0160 .009 \quad 0.000 \quad 0.0010 .0010 .0170 .0140 .014$ 30.3280 .1450 .0980 .0570 .0120 .0120 .0150 .1970 .1200 .106 $40.3610 .245 \quad 0.109 \quad 0.090 \quad 0.039 \quad 0.015 \quad 0.015 \quad 0.233 \quad 0.205 \quad 0.228$ $\begin{array}{llllllllllll}5 & 0.369 & 0.179 & 0.165 & 0.067 & 0.039 & 0.032 & 0.012 & 0.118 & 0.201 & 0.299\end{array}$ 60.2520 .1510 .0810 .0960 .0250 .0170 .0210 .0890 .0930 .333 70.2280 .0980 .1260 .0530 .0460 .0180 .0110 .1850 .0960 .179 $8 \quad 0.228 \quad 0.098 \quad 0.126 \quad 0.053 \quad 0.046 \quad 0.018 \quad 0.011 \quad 0.185 \quad 0.096 \quad 0.179$

| XSA population number (Thousand) age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 2 | 3 |  | 5 | 6 | 7 | 8 |  |
| 2012 | 872 | 557 | 1078 | 935 | 547 | 269 | 593 |  |
| 2013 | 639 | 773 | 363 | 680 | 585 | 384 | 926 |  |
| 2014 | 775 | 556 | 605 | 257 | 514 | 455 | 499 |  |
| 2015 | 1714 | 690 | 457 | 491 | 197 | 429 | 1510 |  |
| 2016 | 3931 | 1537 | 590 | 378 | 416 | 162 | 763 |  |
| 2017 | 1755 | 3556 | 1374 | 513 | 329 | 367 | 795 |  |
| 2018 | 3596 | 1586 | 3178 | 1224 | 450 | 293 | 975 |  |
| 2019 | 2964 | 3250 | 1414 | 2832 | 1095 | 399 | 510 |  |
| 2020 | 7051 | 2636 | 2414 | 1014 | 2277 | 906 | 1489 |  |
| 2021 | 1405 | 6289 | 2116 | 1780 |  | 1878 | 1445 |  |
| Estimated population abundance at 1st Jan 2022 age |  |  |  |  |  |  |  |  |
| year | 2 | 3 | 4 |  | 6 |  |  |  |
| 2022 | 0125 | 54511 | 1815 | 2511 | 95487 | 1422 |  |  |

Fleet: UK (E\&W)-BTS-Q3
Log catchability residuals.

| age | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.035 | 0.018 | 0.403 | 0.501 | -0.059 | -0.282 | 0.154 | 0.168 | -0.288 | 0.090 | 0.439 | -0.156 | -0.007 | -0.050 | -0.910 | 0.130 | 0.032 | -0.006 |
| 3 | 0.574 | 0.357 | -0.135 | -0.302 | 0.463 | -0.283 | -0.054 | 0.285 | -0.687 | -0.084 | 0.103 | 0.002 | -0.215 | -0.232 | -0.235 | -0.185 | 0.410 | -0.374 |
| 4 | -0.008 | 0.054 | -0.254 | -0.938 | 0.438 | -0.112 | -0.300 | 0.038 | -0.258 | -0.178 | -0.779 | 0.313 | 0.316 | -0.495 | 0.058 | 0.226 | -0.113 | -0.227 |
| 5 | -0.396 | -0.030 | 0.956 | -0.623 | -0.028 | -0.321 | 0.020 | -0.591 | -0.229 | 0.026 | -0.767 | 0.329 | -0.122 | -0.142 | -0.392 | 0.198 | 0.439 | -0.075 |
| 6 | -0.240 | -0.241 | 0.295 | -0.205 | 0.166 | -0.073 | 0.538 | -0.021 | -0.179 | -0.159 | -0.280 | 0.351 | 0.151 | -0.094 | 0.078 | 0.005 | 0.040 | 0.175 |
| 7 | -0.122 | 0.087 | 0.195 | -0.185 | -0.191 | -0.068 | 0.198 | -0.330 | -0.142 | 0.288 | 0.211 | 0.194 | -0.110 | -0.008 | -0.009 | -0.226 | 0.360 | -0.019 |
|  | ar |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |  |  |
| 2 | 0.268 | -0.240 | 0.011 | -0.003 | -0.596 | -0.166 | -0.156 | 0.290 | -0.202 | 0.248 | 0.324 | -0.178 | 0.107 | 0.108 | NA -0. | . 028 |  |  |
| 3 | 0.142 | 0.240 | 0.020 | 0.042 | -0.117 | 0.039 | -0.256 | 0.050 | -0.151 | 0.223 | 0.214 | 0.011 | -0.039 | 0.191 | NA -0. | . 019 |  |  |
| 4 | -0.102 | 0.258 | 0.037 | 0.200 | -0.076 | 0.375 | 0.298 | 0.481 | 0.370 | 0.077 | 0.182 | -0.054 | 0.052 | 0.285 | NA -0. | . 167 |  |  |
| 5 | 0.713 | 0.261 | 0.390 | 0.466 | -0.192 | 0.262 | -0.163 | 0.252 | 0.356 | 0.126 | 0.256 | 0.097 | -0.329 | 0.162 | NA -0. | . 474 |  |  |
| 6 | 0.246 | -0.030 | 0.125 | 0.369 | -0.383 | -0.104 | -0.453 | -0.226 | 0.214 | -0.464 | 0.331 | 0.060 | 0.086 | -0.182 | NA -0. | . 130 |  |  |
|  | -0.198 | -0.02 | -0.197 | 0.077 | 0.517 | 0.004 | 0.15 | 0.305 | 0.15 | -0.17 | 192 | -0.560 | 0.339 | -0.308 | NA | 349 |  |  |

Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time
$\begin{array}{lrrrrrr}2 & 3 & 4 & 5 & 6 & 7\end{array}$
$\begin{array}{lrrrrrr}\text { Mean_Logq } & -7.4328 & -7.7576 & -7.8903 & -7.8903 & -7.8903 & -7.8903 \\ \text { S.E_Logq } & 0.2924 & 0.2924 & 0.2924 & 0.2924 & 0.2924 & 0.2924\end{array}$

Terminal year survivor and $F$ summaries:
Age 2 Year class $=2019$
source
scaledWts survivors yrcls
UK (E\&W)-BTS-Q3 $0.961 \quad 12202019$
fshk $0.039 \quad 24932019$

Age 3 Year class $=2018$
source

|  | scaledWts | survivors | yrcls |
| :--- | ---: | ---: | ---: |
| UK (E\&W)-BTS-Q3 | 0.957 | 5024 | 2018 |
| fshk | 0.043 | 7762 | 2018 |

Age 4 Year class $=2017$
source

|  | scaledWts | survivors | yrcls |
| :--- | ---: | ---: | ---: |
| UK (E\&W)-BTS-Q3 | 0.943 | 1290 | 2017 |
| fshk | 0.057 | 3653 | 2017 |

Age 5 Year class $=2016$
source

| UK (E\&W)-BTS-Q3 | 0.915 | 7442016 |
| :--- | ---: | ---: | ---: |

fshk $0.085 \quad 49612016$

Age 6 Year class $=2015$
source
scaledWts survivors yrcls UK (E\&W)-BTS-Q3 $0.947 \quad 4282015$ fshk $0.053 \quad 38282015$

Age 7 Year class $=2014$
source
$\begin{array}{lrrr} & \text { scaledWts } & \text { survivors } \text { yrcls } \\ \text { UK (E\&W)-BTS-Q3 } & 0.954 & 1003 & 2014\end{array}$
fshk $0.046 \quad 8352014$

Table 31.10. Sol.27.7a - Fishing mortality.

| year/age | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Fbar(4-7) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 0.008 | 0.12 | 0.296 | 0.444 | 0.429 | 0.391 | 0.391 | 0.39 |
| 1971 | 0.012 | 0.148 | 0.399 | 0.554 | 0.367 | 0.441 | 0.441 | 0.44 |
| 1972 | 0.01 | 0.081 | 0.352 | 0.506 | 0.493 | 0.452 | 0.452 | 0.451 |
| 1973 | 0.03 | 0.144 | 0.362 | 0.439 | 0.487 | 0.431 | 0.431 | 0.43 |
| 1974 | 0.004 | 0.085 | 0.316 | 0.472 | 0.544 | 0.445 | 0.445 | 0.444 |
| 1975 | 0.042 | 0.157 | 0.303 | 0.485 | 0.397 | 0.396 | 0.396 | 0.395 |
| 1976 | 0.008 | 0.07 | 0.419 | 0.482 | 0.379 | 0.428 | 0.428 | 0.427 |
| 1977 | 0.015 | 0.135 | 0.326 | 0.407 | 0.375 | 0.37 | 0.37 | 0.37 |
| 1978 | 0.008 | 0.074 | 0.287 | 0.404 | 0.382 | 0.358 | 0.358 | 0.358 |
| 1979 | 0.013 | 0.143 | 0.364 | 0.632 | 0.426 | 0.476 | 0.476 | 0.475 |
| 1980 | 0.04 | 0.134 | 0.393 | 0.567 | 0.949 | 0.639 | 0.639 | 0.637 |
| 1981 | 0.017 | 0.149 | 0.329 | 0.511 | 0.603 | 0.483 | 0.483 | 0.481 |
| 1982 | 0.003 | 0.095 | 0.477 | 0.408 | 0.437 | 0.442 | 0.442 | 0.441 |
| 1983 | 0.007 | 0.081 | 0.384 | 0.532 | 0.392 | 0.438 | 0.438 | 0.436 |
| 1984 | 0.045 | 0.181 | 0.243 | 0.38 | 0.433 | 0.353 | 0.353 | 0.352 |
| 1985 | 0.01 | 0.134 | 0.361 | 0.321 | 0.328 | 0.337 | 0.337 | 0.337 |
| 1986 | 0.006 | 0.276 | 0.422 | 0.553 | 0.338 | 0.439 | 0.439 | 0.438 |
| 1987 | 0.059 | 0.181 | 0.566 | 0.763 | 1.26 | 0.868 | 0.868 | 0.864 |
| 1988 | 0.01 | 0.172 | 0.371 | 0.563 | 0.734 | 1.112 | 1.112 | 0.695 |
| 1989 | 0.044 | 0.3 | 0.451 | 0.539 | 0.633 | 0.704 | 0.704 | 0.582 |
| 1990 | 0.113 | 0.4 | 0.682 | 0.666 | 0.616 | 0.676 | 0.676 | 0.66 |
| 1991 | 0.116 | 0.353 | 0.487 | 0.399 | 0.503 | 0.645 | 0.645 | 0.508 |
| 1992 | 0.08 | 0.289 | 0.413 | 0.613 | 0.42 | 0.69 | 0.69 | 0.534 |
| 1993 | 0.014 | 0.149 | 0.355 | 0.422 | 0.548 | 0.925 | 0.925 | 0.563 |
| 1994 | 0.025 | 0.296 | 0.43 | 0.476 | 0.527 | 0.559 | 0.559 | 0.498 |
| 1995 | 0.072 | 0.235 | 0.527 | 0.526 | 0.563 | 0.428 | 0.428 | 0.511 |
| 1996 | 0.026 | 0.346 | 0.475 | 0.5 | 0.591 | 0.547 | 0.547 | 0.528 |
| 1997 | 0.105 | 0.418 | 0.62 | 0.515 | 0.503 | 0.765 | 0.765 | 0.601 |
| 1998 | 0.026 | 0.311 | 0.468 | 0.716 | 0.4 | 0.412 | 0.412 | 0.499 |
| 1999 | 0.062 | 0.206 | 0.348 | 0.546 | 0.734 | 0.268 | 0.268 | 0.474 |
| 2000 | 0.027 | 0.241 | 0.241 | 0.247 | 0.519 | 0.847 | 0.847 | 0.464 |
| 2001 | 0.057 | 0.284 | 0.319 | 0.273 | 0.307 | 0.513 | 0.513 | 0.353 |
| 2002 | 0.069 | 0.29 | 0.527 | 0.469 | 0.325 | 0.22 | 0.22 | 0.385 |
| 2003 | 0.163 | 0.582 | 0.489 | 0.471 | 0.315 | 0.175 | 0.175 | 0.363 |
| 2004 | 0.091 | 0.488 | 0.449 | 0.308 | 0.294 | 0.208 | 0.208 | 0.315 |
| 2005 | 0.212 | 0.459 | 0.703 | 0.686 | 0.416 | 0.425 | 0.425 | 0.557 |
| 2006 | 0.093 | 0.396 | 0.506 | 0.456 | 0.603 | 0.32 | 0.32 | 0.471 |
| 2007 | 0.101 | 0.422 | 0.326 | 0.318 | 0.314 | 0.404 | 0.404 | 0.341 |
| 2008 | 0.055 | 0.279 | 0.372 | 0.274 | 0.311 | 0.308 | 0.308 | 0.316 |


| year/age | 2 | 3 |  | 4 | 5 | 6 | 7 | $8+$ | Fbar(4-7) |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 0.044 | 0.302 | 0.409 | 0.48 | 0.268 | 0.396 | 0.396 | 0.388 |  |
| 2010 | 0.015 | 0.199 | 0.248 | 0.352 | 0.362 | 0.205 | 0.205 | 0.292 |  |
| 2011 | 0.029 | 0.181 | 0.342 | 0.267 | 0.295 | 0.511 | 0.511 | 0.354 |  |
| 2012 | 0.021 | 0.328 | 0.361 | 0.369 | 0.252 | 0.228 | 0.228 | 0.303 |  |
| 2013 | 0.039 | 0.145 | 0.245 | 0.179 | 0.151 | 0.098 | 0.098 | 0.168 |  |
| 2014 | 0.016 | 0.098 | 0.109 | 0.165 | 0.081 | 0.126 | 0.126 | 0.12 |  |
| 2015 | 0.009 | 0.057 | 0.09 | 0.067 | 0.096 | 0.053 | 0.053 | 0.076 |  |
| 2016 | 0 | 0.012 | 0.039 | 0.039 | 0.025 | 0.046 | 0.046 | 0.037 |  |
| 2017 | 0.001 | 0.012 | 0.015 | 0.032 | 0.017 | 0.018 | 0.018 | 0.02 |  |
| 2018 | 0.001 | 0.015 | 0.015 | 0.012 | 0.021 | 0.011 | 0.011 | 0.015 |  |
| 2019 | 0.017 | 0.197 | 0.233 | 0.118 | 0.089 | 0.185 | 0.185 | 0.156 |  |
| 2020 | 0.014 | 0.12 | 0.205 | 0.201 | 0.093 | 0.096 | 0.096 | 0.149 |  |
| 2021 | 0.014 | 0.106 | 0.228 | 0.299 | 0.333 | 0.179 | 0.179 | 0.259 |  |

Table 31.11. Sol.27.7a - Stock numbers-at-age (start of year, in thousands).

| year/age | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 3695 | 8349 | 4145 | 1368 | 4389 | 939 | 8212 | 31097 |
| 1971 | 10177 | 3316 | 6703 | 2791 | 794 | 2585 | 5534 | 31900 |
| 1972 | 3186 | 9101 | 2588 | 4071 | 1451 | 498 | 4321 | 25214 |
| 1973 | 13133 | 2853 | 7595 | 1647 | 2221 | 802 | 3418 | 31667 |
| 1974 | 5870 | 11533 | 2236 | 4784 | 960 | 1234 | 2829 | 29446 |
| 1975 | 6679 | 5288 | 9588 | 1475 | 2699 | 504 | 3220 | 29455 |
| 1976 | 3857 | 5795 | 4087 | 6406 | 822 | 1642 | 2221 | 24830 |
| 1977 | 15772 | 3462 | 4887 | 2432 | 3581 | 509 | 2192 | 32835 |
| 1978 | 9040 | 14060 | 2737 | 3193 | 1464 | 2227 | 2042 | 34763 |
| 1979 | 8847 | 8118 | 11811 | 1859 | 1930 | 905 | 1713 | 35184 |
| 1980 | 5070 | 7902 | 6369 | 7423 | 894 | 1140 | 2536 | 31334 |
| 1981 | 4496 | 4410 | 6257 | 3889 | 3808 | 313 | 2365 | 25538 |
| 1982 | 2460 | 4002 | 3438 | 4074 | 2111 | 1886 | 1163 | 19134 |
| 1983 | 5556 | 2218 | 3292 | 1930 | 2452 | 1234 | 2094 | 18775 |
| 1984 | 15457 | 4992 | 1850 | 2029 | 1026 | 1498 | 1962 | 28815 |
| 1985 | 16223 | 13367 | 3769 | 1313 | 1256 | 602 | 2703 | 39234 |
| 1986 | 23690 | 14533 | 10573 | 2378 | 862 | 819 | 2497 | 55353 |
| 1987 | 3454 | 21302 | 9978 | 6271 | 1238 | 556 | 1453 | 44253 |
| 1988 | 3490 | 2946 | 16091 | 5125 | 2645 | 318 | 800 | 31415 |
| 1989 | 4366 | 3128 | 2244 | 10043 | 2640 | 1149 | 527 | 24096 |
| 1990 | 5547 | 3780 | 2097 | 1293 | 5303 | 1268 | 992 | 20279 |
| 1991 | 12650 | 4482 | 2293 | 959 | 601 | 2592 | 1008 | 24585 |
| 1992 | 4943 | 10195 | 2848 | 1275 | 582 | 329 | 1945 | 22117 |
| 1993 | 6174 | 4128 | 6912 | 1705 | 625 | 346 | 1185 | 21075 |
| 1994 | 5230 | 5508 | 3219 | 4385 | 1011 | 327 | 1282 | 20962 |
| 1995 | 2004 | 4616 | 3706 | 1895 | 2466 | 540 | 1051 | 16278 |
| 1996 | 2493 | 1687 | 3302 | 1980 | 1014 | 1270 | 862 | 12609 |
| 1997 | 8338 | 2199 | 1081 | 1858 | 1087 | 508 | 1039 | 16108 |
| 1998 | 6849 | 6793 | 1310 | 526 | 1004 | 595 | 1312 | 18389 |
| 1999 | 5230 | 6038 | 4503 | 742 | 233 | 609 | 1036 | 18391 |
| 2000 | 6943 | 4446 | 4446 | 2877 | 389 | 101 | 470 | 19672 |
| 2001 | 4543 | 6114 | 3161 | 3160 | 2032 | 209 | 709 | 19927 |
| 2002 | 2328 | 3882 | 4164 | 2078 | 2177 | 1353 | 1153 | 17135 |
| 2003 | 3052 | 1966 | 2628 | 2224 | 1176 | 1423 | 2252 | 14721 |
| 2004 | 3632 | 2346 | 994 | 1459 | 1256 | 776 | 1007 | 11470 |
| 2005 | 2945 | 3002 | 1303 | 574 | 970 | 847 | 1113 | 10754 |
| 2006 | 1318 | 2155 | 1716 | 584 | 262 | 579 | 1450 | 8063 |
| 2007 | 1863 | 1086 | 1313 | 936 | 335 | 130 | 938 | 6600 |


| year/age | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 1949 | 1523 | 644 | 857 | 616 | 221 | 827 | 6638 |
| 2009 | 2263 | 1670 | 1043 | 402 | 590 | 408 | 519 | 6895 |
| 2010 | 1602 | 1961 | 1117 | 627 | 225 | 408 | 894 | 6834 |
| 2011 | 633 | 1428 | 1454 | 789 | 399 | 142 | 319 | 5164 |
| 2012 | 872 | 557 | 1078 | 935 | 547 | 269 | 593 | 4850 |
| 2013 | 639 | 773 | 363 | 680 | 585 | 384 | 926 | 4350 |
| 2014 | 775 | 556 | 605 | 257 | 514 | 455 | 499 | 3662 |
| 2015 | 1714 | 690 | 457 | 491 | 197 | 429 | 1510 | 5489 |
| 2016 | 3931 | 1537 | 590 | 378 | 416 | 162 | 763 | 7776 |
| 2017 | 1755 | 3556 | 1374 | 513 | 329 | 367 | 795 | 8688 |
| 2018 | 3596 | 1586 | 3178 | 1224 | 450 | 293 | 975 | 11302 |
| 2019 | 2964 | 3250 | 1414 | 2832 | 1095 | 399 | 510 | 12463 |
| 2020 | 7051 | 2636 | 2414 | 1014 | 2277 | 906 | 1489 | 17787 |
| 2021 | 1405 | 6289 | 2116 | 1780 | 751 | 1878 | 1445 | 15664 |

Table 31.12. Sol.27.7a - Summary.

| year | recruits | tsb | ssb | landings | Y/ssb | fbar4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 3695 | 7132 | 6436 | 1785 | 0.280 | 0.390 |
| 1971 | 10177 | 7406 | 6222 | 1882 | 0.300 | 0.440 |
| 1972 | 3186 | 5727 | 5011 | 1450 | 0.290 | 0.451 |
| 1973 | 13133 | 6553 | 5123 | 1428 | 0.280 | 0.430 |
| 1974 | 5870 | 6189 | 5068 | 1307 | 0.260 | 0.444 |
| 1975 | 6679 | 6229 | 5359 | 1441 | 0.270 | 0.395 |
| 1976 | 3857 | 5501 | 4889 | 1463 | 0.300 | 0.427 |
| 1977 | 15772 | 5509 | 4490 | 1147 | 0.260 | 0.370 |
| 1978 | 9040 | 6244 | 5092 | 1106 | 0.220 | 0.358 |
| 1979 | 8847 | 6887 | 5684 | 1614 | 0.280 | 0.475 |
| 1980 | 5070 | 6429 | 5513 | 1941 | 0.350 | 0.637 |
| 1981 | 4496 | 5908 | 5165 | 1667 | 0.320 | 0.481 |
| 1982 | 2460 | 4745 | 4330 | 1338 | 0.310 | 0.441 |
| 1983 | 5556 | 4916 | 4095 | 1169 | 0.290 | 0.436 |
| 1984 | 15457 | 6783 | 4601 | 1058 | 0.230 | 0.352 |
| 1985 | 16223 | 7849 | 5635 | 1146 | 0.200 | 0.337 |
| 1986 | 23690 | 9503 | 6944 | 1995 | 0.290 | 0.438 |
| 1987 | 3454 | 8538 | 7148 | 2808 | 0.390 | 0.864 |
| 1988 | 3490 | 5975 | 5498 | 1999 | 0.360 | 0.695 |
| 1989 | 4366 | 5156 | 4609 | 1833 | 0.400 | 0.582 |
| 1990 | 5547 | 4262 | 3603 | 1583 | 0.440 | 0.660 |
| 1991 | 12650 | 4450 | 3165 | 1212 | 0.380 | 0.508 |
| 1992 | 4943 | 4434 | 3430 | 1259 | 0.370 | 0.534 |
| 1993 | 6174 | 3840 | 3212 | 1023 | 0.320 | 0.563 |
| 1994 | 5230 | 4951 | 4023 | 1374 | 0.340 | 0.498 |
| 1995 | 2004 | 3934 | 3493 | 1266 | 0.360 | 0.511 |
| 1996 | 2493 | 3057 | 2689 | 1002 | 0.370 | 0.528 |
| 1997 | 8338 | 3420 | 2480 | 1003 | 0.400 | 0.601 |
| 1998 | 6849 | 4216 | 2998 | 911 | 0.300 | 0.499 |
| 1999 | 5230 | 4287 | 3288 | 863 | 0.260 | 0.474 |
| 2000 | 6943 | 3883 | 3097 | 818 | 0.260 | 0.464 |
| 2001 | 4543 | 4276 | 3531 | 1053 | 0.300 | 0.353 |
| 2002 | 2328 | 3980 | 3539 | 1090 | 0.310 | 0.385 |
| 2003 | 3052 | 3573 | 3174 | 1014 | 0.320 | 0.363 |
| 2004 | 3632 | 2754 | 2274 | 709 | 0.310 | 0.315 |
| 2005 | 2945 | 2493 | 2055 | 855 | 0.420 | 0.557 |
| 2006 | 1318 | 1872 | 1623 | 569 | 0.350 | 0.471 |
| 2007 | 1863 | 1640 | 1383 | 492 | 0.360 | 0.341 |


| year | recruits | tsb | ssb | landings | Y/ssb | fbar4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 1949 | 1571 | 1320 | 332 | 0.250 | 0.316 |
| 2009 | 2263 | 1341 | 1069 | 325 | 0.300 | 0.388 |
| 2010 | 1602 | 1451 | 1190 | 277 | 0.230 | 0.292 |
| 2011 | 633 | 1249 | 1091 | 330 | 0.300 | 0.354 |
| 2012 | 872 | 1283 | 1150 | 298 | 0.260 | 0.303 |
| 2013 | 639 | 1191 | 1081 | 148 | 0.140 | 0.168 |
| 2014 | 775 | 955 | 866 | 99 | 0.110 | 0.120 |
| 2015 | 1714 | 1492 | 1284 | 76 | 0.060 | 0.076 |
| 2016 | 3931 | 1594 | 1189 | 35 | 0.030 | 0.037 |
| 2017 | 1755 | 2193 | 1794 | 34 | 0.020 | 0.020 |
| 2018 | 3596 | 2955 | 2494 | 36 | 0.010 | 0.015 |
| 2019 | 2964 | 3207 | 2703 | 400 | 0.150 | 0.156 |
| 2020 | 7051 | 3997 | 3298 | 404 | 0.120 | 0.149 |
| 2021 | 1405 | 3879 | 3385 | 629 | 0.190 | 0.259 |

Table 31.13. Sole in 7.a - Input for catch forecast and $\mathrm{F}_{\text {MSY }}$ analysis.
Input: $\quad$ F 2022: TAC constraint for 2022 (projected landings 2022: 696 t)

F 2023-2024: mean 19-21 scaled to 2021
Catch and stock weights are mean 19-21
Recruits age 2 in 2022-2024 GM(12-20)

| 2022 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 1937 | 0.1 | 0.38 | 0 | 0 | 0.134 | 0.0174 | 0.18 |
| 3 | 1254 | 0.1 | 0.71 | 0 | 0 | 0.2 | 0.1627 | 0.222 |
| 4 | 5118 | 0.1 | 0.97 | 0 | 0 | 0.247 | 0.2559 | 0.269 |
| 5 | 1525 | 0.1 | 0.98 | 0 | 0 | 0.306 | 0.2377 | 0.328 |
| 6 | 1195 | 0.1 | 1 | 0 | 0 | 0.356 | 0.198 | 0.364 |
| 7 | 487 | 0.1 | 1 | 0 | 0 | 0.385 | 0.1768 | 0.357 |
| +gp | 2515 | 0.1 | 1 | 0 | 0 | 0.404 | 0.1768 | 0.392 |
| Fbar 4-7 |  |  |  |  |  |  | 0.2171 |  |


| 2023 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 1937 | 0.1 | 0.38 | 0 | 0 | 0.134 | 0.0208 | 0.18 |
| 3 | 1723 | 0.1 | 0.71 | 0 | 0 | 0.2 | 0.1944 | 0.222 |
| 4 | 964 | 0.1 | 0.97 | 0 | 0 | 0.247 | 0.3058 | 0.269 |
| 5 | 3585 | 0.1 | 0.98 | 0 | 0 | 0.306 | 0.2841 | 0.328 |
| 6 | 1088 | 0.1 | 1 | 0 | 0 | 0.356 | 0.2367 | 0.364 |
| 7 | 887 | 0.1 | 1 | 0 | 0 | 0.385 | 0.2113 | 0.357 |
| +gp | 2276 | 0.1 | 1 | 0 | 0 | 0.404 | 0.2113 | 0.392 |
| Fbar 4-7 |  |  |  |  |  |  | 0.2595 |  |


| 2023 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 1937 | 0.1 | 0.38 | 0 | 0 | 0.134 | 0.0208 | 0.18 |
| 3 | 1723 | 0.1 | 0.71 | 0 | 0 | 0.2 | 0.1944 | 0.222 |
| 4 | 964 | 0.1 | 0.97 | 0 | 0 | 0.247 | 0.3058 | 0.269 |
| 5 | 3585 | 0.1 | 0.98 | 0 | 0 | 0.306 | 0.2841 | 0.328 |
| 6 | 1088 | 0.1 | 1 | 0 | 0 | 0.356 | 0.2367 | 0.364 |
| 7 | 887 | 0.1 | 1 | 0 | 0 | 0.385 | 0.2113 | 0.357 |
| +gp | 2276 | 0.1 | 1 | 0 | 0 | 0.404 | 0.2113 | 0.392 |
| Fbar 4-7 |  |  |  |  |  |  | 0.2595 |  |

Table 31.14. Sol.27.7a - Management option table.
F 2022: TAC constraint for 2022 (projected landings 2022: 696 t)

F 2023-2024: mean 19-21 scaled to 2021
Catch and stock weights are mean 19-21

Recruits age 2 in 2022-2024 GM(12-20)
$F_{\text {bar }}$ age range: 4-7

| 2022 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| SSB | $\mathrm{F}_{\text {Mult }}$ | $\mathrm{F}_{\text {bar }}$ | Landings |  |
| 3588 | 0.8 | 0.217 | 696 |  |
| 2023 |  |  |  | 2024 |
| SSB | $\mathrm{F}_{\text {Mult }}$ | $\mathrm{F}_{\text {bar }}$ | Landings | SSB |
| 3299 | 0.0 | 0.000 | 0 | 3674 |
| 3299 | 0.1 | 0.026 | 79 | 3593 |
| 3299 | 0.2 | 0.052 | 157 | 3514 |
| 3299 | 0.3 | 0.078 | 232 | 3437 |
| 3299 | 0.4 | 0.104 | 306 | 3361 |
| 3299 | 0.5 | 0.130 | 378 | 3288 |
| 3299 | 0.6 | 0.156 | 448 | 3217 |
| 3299 | 0.7 | 0.182 | 517 | 3147 |
| 3299 | 0.8 | 0.208 | 584 | 3079 |
| 3299 | 0.9 | 0.234 | 649 | 3012 |
| 3299 | 1.0 | 0.259 | 712 | 2948 |
| 3299 | 1.1 | 0.285 | 775 | 2885 |
| 3299 | 1.2 | 0.311 | 835 | 2823 |
| 3299 | 1.3 | 0.337 | 895 | 2763 |
| 3299 | 1.4 | 0.363 | 952 | 2704 |
| 3299 | 1.5 | 0.389 | 1009 | 2647 |
| 3299 | 1.6 | 0.415 | 1064 | 2591 |
| 3299 | 1.7 | 0.441 | 1118 | 2537 |
| 3299 | 1.8 | 0.467 | 1170 | 2483 |
| 3299 | 1.9 | 0.493 | 1222 | 2432 |
| 3299 | 2.0 | 0.519 | 1272 | 2381 |

Input units are thousands and $\mathbf{k g}$ - output in tonnes.

| Basis | Catch_ <br> 2023 | $\begin{aligned} & \text { Landings_ } \\ & 2023 \end{aligned}$ | $\begin{aligned} & \text { Discards_ } \\ & 2023 \end{aligned}$ | Fbar_ $2023$ | $\begin{aligned} & \text { SSB_ } \\ & 2024 \end{aligned}$ | SSB_ <br> change | TAC_ <br> change | Advice_ change |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{F}_{\text {MSY_rescal }}$ | 605 | 535 | 70 | 0 | 3129 | -5.2 | -23 | -23 |
| FMSY _lower_rescal | 492 | 435 | 57 | 0.151 | 3230 | -2.1 | -37 | -37 |
| $\mathrm{F}_{0}$ | 0 | 0 | 0 | 0 | 3674 | 11.4 | -100 | -100 |
| $\mathrm{F}_{\mathrm{pa}}$ | 695 | 615 | 81 | 0.22 | 3047 | -7.6 | -11.6 | -11.6 |
| $F_{\text {lim }}$ | 888 | 786 | 103 | 0.29 | 2874 | -12.9 | 12.9 | 12.9 |
| $\mathrm{B}_{\text {lim }}$ | 1305 | 1154 | 151 | 0.46 | 2500 | -24 | 66 | 66 |
| $\mathrm{B}_{\mathrm{pa}}$ | 192 | 170 | 22 | 0.056 | 3500 | 6.1 | -76 | -76 |
| $F_{\text {int }}$ | 687 | 608 | 80 | 0.22 | 3054 | -7.4 | -12.7 | -12.7 |
| $\mathrm{B}_{\text {intplus1 }}$ | 415 | 367 | 48 | 0.126 | 3299 | 0 | -47 | -47 |
| $\mathrm{F}_{\text {MSY _ }}$ lower | 520 | 460 | 60 | 0.16 | 3205 | -2.9 | -34 | -34 |
| $\mathrm{F}_{\text {MSY }}$ | 638 | 564 | 74 | 0.2 | 3099 | -6.1 | -18.9 | -18.9 |
| $\mathrm{F}_{\text {MSY _ }}$ upper | 752 | 665 | 87 | 0.24 | 2996 | -9.2 | -4.5 | -4.5 |

Table 31.15. Sol.27.7a - Detailed output of the short-term forecast (MSY approach: $F=F_{\text {MSY }} \times$ SSB $_{2023} /$ MSY $_{\text {trigger }}$ ). Input units are thousands and $\mathbf{k g}$ - output in tonnes.

| Year: | 2022 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 2 | 0.01736 | 32 | 6 | 1937 | 259 | 736 | 98 |
| 3 | 0.16266 | 179 | 40 | 1254 | 251 | 890 | 178 |
| 4 | 0.25588 | 1102 | 296 | 5118 | 1262 | 4964 | 1225 |
| 5 | 0.23766 | 308 | 101 | 1525 | 467 | 1494 | 458 |
| 6 | 0.19804 | 205 | 74 | 1195 | 426 | 1195 | 426 |
| 7 | 0.17678 | 75 | 27 | 487 | 187 | 487 | 187 |
| 8 | 0.17678 | 388 | 152 | 2515 | 1017 | 2515 | 1017 |
| Total |  | 2289 | 696 | 14030 | 3869 | 12282 | 3588 |
| $\mathrm{F}_{\text {bar }} 4-7$ | 0.21709 |  |  |  |  |  |  |
| Year: | 2023 |  |  |  |  |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 2 | 0.01508 | 28 | 5 | 1937 | 259 | 736 | 98 |
| 3 | 0.14125 | 216 | 48 | 1723 | 345 | 1223 | 245 |
| 4 | 0.22219 | 183 | 49 | 964 | 238 | 935 | 231 |
| 5 | 0.20638 | 637 | 209 | 3585 | 1098 | 3514 | 1076 |
| 6 | 0.17197 | 164 | 60 | 1088 | 388 | 1088 | 388 |
| 7 | 0.15351 | 120 | 43 | 887 | 341 | 887 | 341 |
| 8 | 0.15351 | 309 | 121 | 2276 | 920 | 2276 | 920 |
| Total |  | 1657 | 535 | 12460 | 3588 | 10659 | 3299 |
| $\mathrm{F}_{\text {bar }} 4-7$ | 0.18851 |  |  |  |  |  |  |


| Year: | 2024 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 2 | 0.01508 | 28 | 5 | 1937 | 259 | 736 | 98 |
| 3 | 0.14125 | 217 | 48 | 1727 | 345 | 1226 | 245 |
| 4 | 0.22219 | 257 | 69 | 1353 | 334 | 1313 | 324 |
| 5 | 0.20638 | 124 | 41 | 699 | 214 | 685 | 210 |
| 6 | 0.17197 | 397 | 145 | 2639 | 940 | 2639 | 940 |
| 7 | 0.15351 | 112 | 40 | 829 | 319 | 829 | 319 |
| 8 | 0.15351 | 333 | 130 | 2455 | 992 | 2455 | 992 |
| Total |  | 1468 | 478 | 11638 | 3404 | 9882 | 3129 |
| $\mathrm{F}_{\mathrm{bar}} 4-7$ | 0.18851 |  |  |  |  |  |  |



Figure 31.1a. Sol.27.7a - Age composition of landings.


Figure 31.2. Sol.27.7a - Standardized landings proportion.


Catch category

- DIS
- LAN

Figure 31.3.- Sol.27.7a - BE Length distributions of discarded and retained fish from discard sampling studies (Beam trawl).


Figure 31.4. Figure 31.4 Sole in 7.a - Mean standardised LPUE (kg/100 Km fished) for the UK (E\&W) September beamtrawl survey (UK(E\&W)-BTS-Q3).

Mean standardised effort

fleet
$\rightarrow$ BE-CBT
$\rightarrow$ IR-CBT
-- UK-CBT-new
$\rightarrow$ UK-CBT-arig

Figure 31.5b. Sole in 7.a - Mean standardised effort and LPUE for the commercial otter trawl fleets.
Mean standardised Ipue


## fleet

- BE-CBT
$\rightarrow$ IR-CBT
-- UK-CBT-new
-- UK-CBT-orig

Figure 31.5. Sole in 7.a - Mean standardised effort and LPUE for the commercial beam trawl fleets.
LPUE: BE-CBT and IR-CBT: Kg/hr; UK-CBT-new: Kg/day; UK-CBT-orig: Kg/000'hr fished (GRT corrected $>40$ vessels).

EFFORT: BE-CBT: 000' hours fishing; IR-CBT: 000'hours; UK-CBT-new: days fished; UK-CBTorig: 000'hours fished (GRT corrected $>40$ ' vessels).

Mean standardised effort

fleet
$\rightarrow$ IR-COT
$\rightarrow$ UK-COT-new
$\rightarrow$ UK-COT-orig

Mean standardised Ipue

fleet
$\rightarrow$ IR-COT

Figure 31.6. Sole in 7.a - Mean standardised effort and LPUE for the commercial otter trawl fleets.
LPUE: IR-COT: Kg/hr; UK-COT-new: Kg/day; UK-COT-orig: Kg/000'hr fished (GRT corrected $>40$ ' vessels).

EFFORT: IR-CBT: 000'hours; UK-CBT-new: days fished; UK-CBT-orig: 000'hours fished (GRT corrected $>40^{\prime}$ vessels).

UK (E\&W)-BTS-Q3


Figure 31.7. Sol.27.7.a - Mean-standardised indices.


Figure 31.8. Sol.27.7a - Consistency plot UK(E\&W)-BTS-Q3 survey.


Figure 31.9. Sol.27.7a - Log catchability residual plot - Final XSA.


Figure 31.10. Sol.27.7a - Summary plots.


Figure 31.11. Sol.27.7a - Retrospective XSA analysis (shrinkage SE=1.5).


Figure 31.12. - Sol.27.7a - comparison with last year's assessment.


Figure 31.13. Sol.27.7a - Options for the intermediate year in the short-term forecast.


Figure 31.14. Sol.27.7a - Comparison of international TAC, catch and landings.


Figure 31.15. Sol.27.7a - Output for the short-term forecast under the MSY approach (MSY approach: $F=F_{\text {MSY }} \times$ SSB $_{2023} /$ MSY $_{\text {trigger }}=0.189$ ).

Yield(landings) 2023


SSB 2024


Figure 31.16. Sol.27.7a - Year-class sources and contributions for the short-term forecast (MSY approach: $F=F_{\text {MSY }} \times$ SSB $_{2023} /$ MSY $_{\text {trigger }}=0.189$ ).


Figure 31.17. Sol.27.7a - Stock-recruitment plot.

# 32 Sole (Solea solea) in Division 7.e (western English Channel) 

## Type of assessment in 2021

Last year's assessment report is available at:
https://www.ices.dk/sites/pub/Publication\ Reports/Expert\ Group\ Report/Fisheries\ Resources\ Steering\ Group/2021/WGCSE publication\%20with\%20multiple\%20files/WGCSE2021 32\%20Sole\%207e.pdf

## ICES advice applicable to 2022

Last year's advice is available at https://doi.org/10.17895/ices.advice. 7862 and stated:
ICES advises that when the MSY approach is applied, catches in 2022 should be no more than 1810 tonnes.

### 32.1 Impact of the COVID-19 pandemic

The sole in Division 7.e stock, its fishery, and data sampling were largely unaffected by the implications of the COVID-19 pandemic in 2021.

### 32.2 ICES Transparent Assessment Framework

The Division 7.e sole stock is included in the ICES Transparent Assessment Framework (TAF, https://taf.ices.dk, https://github.com/ices-taf). All WGCSE assessments since 2018 are available from the ICES TAF GitHub page (https://github.com/ices-taf/2018 sol.27.7e, https://github.com/ices-taf/2019 sol.27.7e, https://github.com/ices-taf/2020 sol.27.7e assessment; please note, access to these repositories is so far restricted to ICES and members of WGCSE). The current WGCSE 2022 assessment is available from https://github.com/icestaf/2022 sol.27.7e assessment.

All changes since last year's assessment can be accessed with the following link: https://github.com/ices-taf/2022 sol.27.7e assessment/compare/8331f09...main.

The TAF repository includes all input data, R scripts for processing data, preparing and running the stock assessment and forecast, and scripts for creating all figures and tables presented in this report. This repository also contains documentation on how to reproduce the WGCSE assessment for sole.

### 32.3 General

## Stock description and management units

The TAC specified for ICES Division 7.e is consistent with the assessment area.
Official national landings data as reported to ICES and the landings estimates as used by the Working Group are given in Table 32.1.

Official landings in 2021 were 1395 t, a $28 \%$ undershoot of the 2021 TAC ( 1925 t ).

The TAC and the national quotas by country for 2021

| Species | Sole <br> Solea solea | Zone: | 7e <br> (Sol/07E.) |
| :--- | :--- | :--- | :--- |
| Belgium | 63 | Analytical TAC |  |
| France | 671 |  |  |
| Union | 733 |  |  |
| United Kingdom | 1175 |  |  |
| TAC | 1925 |  |  |

(Source: Council Regulation (EU) 2021/1239, EU, 2021).
The TAC and the national quotas by country for 2022

| Species | Sole <br> Solea solea |  |
| :--- | :---: | :--- |
| Belgium | Zone: | 7e <br> (Sol/07E.) |
| France | 59 | Analytical TAC |
| Union | 631 |  |
| United Kingdom | 690 | 1111 |
| TAC | 1810 |  |

(Source: Council Regulation (EU) 2022/515, EU, 2022).

Maximum number of days a vessel may be present within the area by category of regulated gear per year for 2022

| Regulated gear | Maximum number of days |  |
| :--- | :--- | :--- |
| Beam trawls of mesh size $\geq 80 \mathrm{~mm}$ | BE | 176 |
|  | FR | 188 |
| Static nets with mesh size $\leq 220 \mathrm{~mm}$ | BE | 176 |
|  | FR | 191 |

(Source: Council Regulation (EU) 2022/515, ANNEX III, EU, 2022).

## Landing obligation

As of 2020, the EU landing obligation fully applied to sole in Division 7.e. However, a de minimis exemption allows up to $3 \%$ of total annual catches to be discarded for trammel and gillnets and beam trawls with mesh size $80-119 \mathrm{~mm}$ with a Flemish panel (Commission Delegated Regulation (EU) 2020/2015, EU, 2020).

A landing obligation also applies in UK waters and includes a de minimis exemption for sole for trammel and gillnets (MMO, 2020a) and beam trawls with a Flemish panel (MMO, 2020b).

However, the UK landing obligation specifies the de minimis exemption as a "small percentage of the total catch" without specifying a value.

The EU landing obligation was phased in between 2016-2019 (Commission Delegated Regulations (EU) 2015/2438, 2016/2375, 2018/46, EU, 2015, 2016, 2018). During the phasing in, the landing obligation applied to all catches of sole in 7.e with trammel and gillnets (gear codes GNS, GN, GND, GNC, GTN, GTR, GEN) and all beam trawls. However, a de minimis exemption applied, allowing up to $3 \%$ discards of annual catches for all trammel and gillnets and for beam trawls with a mesh size of 80-199 mm with increased selectivity. In 2016, the first year of the application, the landing obligation applied only to vessels for which the total landings consisted of more than $10 \%$ sole during two reference years (2013 and 2014, Commission Delegated Regulation (EU) 2015/2438, EU, 2015). This threshold was tightened for 2017, and the landing obligation applied to vessels landing more than 5\% in the reference years 2014 and 2015 (Commission Delegated Regulation (EU) 2016/2375, EU, 2016). Subsequently, this restriction was lifted altogether, and for 2018 (Commission Delegated Regulation (EU) 2018/46, EU, 2018), the landing obligation applied to all vessels using trammel and gillnets and beam trawls, as described above.

Given the low discards observed in the fishery, the landing obligation is unlikely to impact this stock or the advice significantly.

### 32.4 Data

## InterCatch

International catch data are collated using the ICES InterCatch platform. For 2021, data for Belgium, France, Ireland, the Netherlands and the United Kingdom (England, Scotland and the Channel Island Guernsey) were uploaded into InterCatch (Figures 32.1 and 32.2). All submitted age samples are presented in Figure 32.8 and length samples in Figure 32.9. The raising procedure is described in the Stock Annex.

## Landings

Landings of sole in Division 7.e were below 500 t at the beginning of the time-series in the 1970s. Subsequently, landings increased and stayed around 1500 t in the 1980 s and have been around 1000 t in the 1990s and 2000s (Table 32.1). The landings dropped in the late 2000s below 750 t and increased since 2015 to 1392 t in 2021.

The UK, France and Belgium provided age-structured landings samples in InterCatch (Figure 32.8).

Total international landings numbers-at-age (Table 32.2 and Figure 32.5) and landings and stock weights-at-age (Tables 32.3 and 32.4 and Figure 32.6), as used in the assessment, were derived following the procedures outlined in the Stock Annex.

The fleets for which age distributions were submitted accounted for $85.1 \%$ of the 2021 total international landings, based on the InterCatch level (year, country, fleet, and quarter), up from 84.8\% for 2020 and $77.8 \%$ for 2019.

## Discards

Discards for this stock are very low and not included in the assessment.

For 2021, discards estimates were provided by Belgium, France, Ireland (0 discards), and the UK for some fleets in InterCatch based on discard sampling. Discard age samples were only provided for the French OTB_DEF_70-99 Q2 and UK TBB_DEF_70-99 Q4.

Discards data are only available from InterCatch for the years 2012-2021. In general, the discard rates are low (Figure 32.3). A higher discard rate was observed in 2015, attributed to high discards from the multirig otter trawl (mesh size $90-99 \mathrm{~mm}$ ) fleet. The three-year average (20192021) discard rate is $0.44 \%$. This reduction in the discard rate might be linked to introducing the landing obligation in 2016.

The discard rate by fleet and country is shown in Figure 32.4 (shown are only discards submitted to InterCatch).

No discard information is included in the assessment, given that it is currently not possible to provide discard estimates for the entire time-series. Nevertheless, excluding discard estimates from the assessment is unlikely to have any major impact on the perception of stock status given the minor scale of the problem.

## Revisions

No revisions to previous years were submitted.

## Biological data

Natural mortality was assumed to be constant over all ages and years at 0.1 . The maturity ogive from divisions 7.f and 7.g was used following the procedures outlined in the Stock Annex and adopted in previous assessments.

In agreement with the Stock Annex, stock and catch weights-at-age were derived by fitting a second-degree polynomial model to the raw landings weights-at-age extracted from InterCatch (Figure 32.7). For 2021 data, the youngest age for which data (catch numbers and weights) were provided was age 1.

## Survey indices

Abundance estimates derived from the surveys as used in the assessment are given in Table 32.6 and shown in Figures 32.10, 32.11, 32.12 and 32.13, and internal consistencies in Figures 32.14, 32.15 and 32.16. In general, cohort tracking and internal consistency are better in the commercial tuning fleets and less pronounced in the scientific surveys.

## The UK-FSP survey

The UK Western Channel sole and plaice survey (previously called Fisheries Science Partnership survey; UK-FSP, quarter 3, ICES survey code B4381, Burt et al., 2022) conducted another survey of sole and plaice abundance in the Western English Channel in 2021. The survey uses two 4 m beam trawls with 80 mm nominal codend mesh and focuses on the area around the English coast. 83 out of 90 tows were completed in 2021. 339 sole otoliths were collected for ageing in 2021.

Catch rates are reported standardised as numbers per hour per meter of beam length. The results indicate that sole continues to be widespread in the area (Figure 32.17) and that many cohorts contribute to the stock. The total CPUE increased since 2016 but dropped slightly in 2020 and 2021. The index is mainly driven by ages 3,4 , and 5 . The internal consistency in the survey is good for ages $3+$. Some year and cohort effects are visible.

## The Q1SWBeam survey

The second survey used for sole is the Quarter 1 South West Beam trawl (Q1SWBeam, also called Q1SWECOS, ICES survey code B2732), which started in 2006. This survey deploys two 4 m beam trawls and uses a fully random stratified approach. In contrast to the FPS survey, the Q1SWBeam covers the entire western English Channel and, if conditions permit, adjacent areas. In 2021, 78 out of 81 tows were completed in the western English Channel. The landings per unit of effort (LPUE) numbers-at-age as well as aggregated over all ages are variable without particular trends or patterns, and internal consistency is mediocre.

Sole are caught in the entire western English Channel with higher numbers along the English coast.

In 2020, this survey was delayed due to disruptions of the COVID pandemic but returned to its normal schedule in 2021.

## Commercial fleets effort and Ipue

Two commercial tuning series from the UK are used (commercial beam trawl UK-CBT and commercial otter trawl UK-COT).

Effort for under 24 m UK beam trawlers in days fished steadily increased from 1992, and reached the highest levels on record in 2012 and stayed around this level until the end of the time-series (Figure 32.10). Currently, the effort is well above the long-term average. In contrast, the effort for over 24 m UK beam trawlers increased from 1992 to 2004 and then decreased to below the average of the time-series, reaching a minimum in 2013. Since then, the effort increased again slightly and is currently around the long-term average. When the effort of all UK beam trawl vessels is combined, the effort stayed almost constant since the early 2000s.
UK otter trawl (UK-COT) effort has been in continual decline since the early 1970s and was at the lowest levels on record in 2015. For 2016, this fleet reported zero effort and landings. This could be explained by a shift in the size of fishing vessels to smaller vessels. Since 2017, a new database is being used for recording, but the data are not consistent with historical data and are therefore not used in the stock assessment.

Age-disaggregated commercial abundance indices for the UK-CBT-late (UK-CBT values from 2003 onwards) and UK-COT fleets as used in the assessment are given in Tables 32.5 and 32.6 and plotted in Figures 32.10-32.13.

### 32.5 Stock assessment

Model used: Extended Survivors Analysis (XSA) as outlined in the Stock Annex by IBPWCFlat2 2015.

Software used: FLR - FLXSA.
Model options chosen: Data included in the assessment were identical to previous years, apart from one additional data year.

Assessment input data characteristics: catch numbers-at-age excluding discards and four tuning fleets (two fishery-independent surveys: UK-FSP and Q1SWBeam; and two commercial LPUE time-series: UK-CBT-late and UK-COT).

## Data screening

Data screening procedures identified no major anomalies in the catch numbers-at-age, weights or tuning information used in the 2022 assessment.
The landings numbers-at-age 3 were exceptionally high in 2017 but returned to usual levels in 2018 (Figure 32.5). This anomaly was evident in age samples from the UK and France and various fleets (see WGCSE 2018 report), i.e. does not seem to be a sampling issue. Another strong cohort appeared in the landings-at-age 3 in 2019, and is tracking well as age 4 in 2020 and age 5 in 2021, and was also visible in the two survey indices and the commercial beam trawl LPUE in 2021.

Tuning information consisted of four fleets: two UK commercial time-series (UK-CBT-late and UK-COT) and two UK standardised research surveys (UK-FSP and Q1SWBeam).

The UK commercial otter trawl fleet (UK-COT) reported zero effort in 2016. Therefore, there is no LPUE value for this fleet for 2016. Consequently, this tuning index only influences the assessment up to and including 2015.

Details of the derivation of the tuning fleets are presented in the Stock Annex. The tuning information available for this assessment is shown in Table 32.6.

## Final update assessment

The working group fitted the XSA model developed by WKFLAT 2012 (ICES, 2012) using the updated assessment settings agreed at IBPWCFlat2 (ICES, 2015).

The XSA assessment settings used at the last three working groups are shown in the table below, and more historical settings have been included in the Stock Annex.

|  | WGCSE 2020 | WGCSE 2021 | WGCSE 2021 |
| :---: | :---: | :---: | :---: |
| Assessment age range | 2-12+ | 2-12+ | 2-12+ |
| $F_{\text {bar }}$ age range | F(3-9) | $F(3-9)$ | $F(3-9)$ |
| Assessment method | XSA | XSA | XSA |
| Tuning Fleets: |  |  |  |
| Q1SWBeam | 2006-2019 | 2006-2020 | 2006-2021 |
|  | Ages 2-11 <br> (non-offset) | Ages 2-11 <br> (non-offset) | Ages 2-11 <br> (non-offset) |
| UK-FSP | 20014-2019 | 20014-2020 | 20014-2021 |
|  | Ages 2-11 | Ages 2-11 | Ages 2-11 |
| UK combined beam (late) | 2003-2019 | 2003-2020 | 2003-2021 |
|  | Ages 3-11 | Ages 3-11 | Ages 3-11 |
| UK otter trawl | 1988-2016 | 1988-2016 | 1988-2016 |
|  | Ages 3-11 | Ages 3-11 | Ages 3-11 |
| Time taper | Yes | Yes | Yes |
| Power model | Tricubic | Tricubic | Tricubic |
| Taper range | 15 years | 15 years | 15 years |
| P shrinkage | No | No | No |
| Q plateau age | 7 | 7 | 7 |
| F shrinkage S.E | 0.5 | 0.5 | 0.5 |
| Number of years | 3 | 3 | 3 |
| Number of ages | 5 | 5 | 5 |
| Fleet S.E. | 0.4 | 0.4 | 0.4 |

Figure 32.18 shows the results from the final XSA model fit, Figure 32.19 the model residuals, Figure 32.20 a comparison of the current assessment with last years' assessments, Figure 32.21 XSA survivor weightings for the last two years and Figure 32.22 a five-year retrospective.

The survey residuals show relatively large values in earlier years. This is an expected feature of the XSA assessment, which includes a taper range of 15 years. This means that older survey observations are down-weighted and any observations 15 years or older are not used in the assessment.

A Mohn's rho analysis was conducted based on the XSA stock assessment results, i.e. the last data year (2021) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 3-9) | recruitment |
| :--- | :---: | :---: | :---: |
| Mohn's rho value | -0.091 | 0.157 | 0.026 |

The Mohn's rho values for this assessment are well within the ICES WKFORBIAS thresholds $(+0.20,-0.15)$, i.e. the current assessment indicates sufficient consistency for advice purposes. The retrospective has been increasing slightly in recent years; however, this does not cause concerns because the SSB is usually underestimated while the fishing mortality is overestimated, i.e. the retrospective pattern is likely only to increase precaution of the assessment and advice.

XSA diagnostic of the final assessment are presented in Table 32.7, stock numbers-at-age in Table 32.8, fishing mortalities-at-age in Table 32.9 and an assessment summary in Table 32.10.

## Consistency of the stock assessment

The comparison of historical stock assessment results (historical retro, Figure 32.20) and the analytical retrospective analysis (Figure 32.22) show slightly different retrospective patterns. It is worth noting that the historical comparison (Figure 32.22) shows the assessment results (including short-term forecast assumptions for the intermediate year) from conducting the stock assessment in the corresponding years. These values are stored in an ICES database and not updated afterwards. On the other hand, the analytical retrospective analysis (Figure 32.22) is conducted with the most recent version of the input data, and its retrospective runs are also based on these most recent (possibly updated or revised) data and only removing data years from the end.

The differences between the historical and analytical retro can be explained through revisions of historical input data over the years, namely:

- Q1SWBeam: The scientific Q1SWBeam survey was revised prior to WGCSE 2020 because of a quality control process of the survey data and data processing. This is described in the WGCSE 2020 report. This has led to some changes in historical data, which means all historical assessment results prior to 2020 are based on slightly different historical Q1SWBeam data compared to the assessment afterwards.
- FSP: The FSP survey index values used in the assessment are a product of a model fitting to the raw data. This means that the entire time-series is updated every year, including historical values.

Furthermore, the terminal year of the historical retro includes assumptions for the intermediate year. In the following year's assessment, the observed perception of the fishery can be different from that assumed in the previous year.

## State of the stock

Stock trends are shown in Table 32.10 and plotted in Figure 32.19. The stock is in a desirable state, both in terms of spawning-stock biomass and fishing mortality.

SSB is estimated to have increased between 1972 and 1980 following successive strong recruitment events. Subsequently, SSB declined from 1980 to 1993 and remained relatively stable until 2008. After this period, SSB increased and is currently well above MSY $B_{\text {trigger. }}$

The base level of recruitment has remained relatively stable throughout the time-series, fluctuating without a major temporal trend at around 4 million recruits. Recruitment variability has decreased since 1991. In recent years, recruitment has been variable again, with very high recruitment estimates in 2018 and 2020 but a very low estimate in 2021.

Fishing mortality was relatively stable at a low level between 1969 and 1978, after which it increased sharply until 1983 and fluctuated at a higher level before peaking briefly in 1989-1990. After a period of temporal variability, F decreased abruptly to below the Fmsy target of 0.29 in 2009. Since then, F has remained below this level but has been increasing again and was just below Fmsy in 2021.

The age structure of sole in 7.e continues to be more extended than other sole stocks in European waters, implying low mortality rates, with the plus group at age 12 containing a relatively high proportion of the catches and including some individuals aged 33-38 in recent years.

### 32.6 Short-term projections

## Forecast assumptions

Figure 32.23 shows three different targets for the intermediate year: status quo ( $\mathrm{F}=\mathrm{F}_{2021}$ ), average ( $\mathrm{F}=\mathrm{F}$ average 2019-2021), and TAC. F estimates decreased slightly in 2020 compared to 2019 but increased in 2021.

Landings have been below the international TAC and the advised catch in previous years (Figure 32.24). The catch advice for 2022 ( 1810 t ) is substantially higher than recent landings, and reaching this level would mean an increase in landings of $30 \%$ compared to 2021 . However, this appears unlikely given the recent trend in landings, fleet capacity and legal limitations on fishing effort.
$\mathrm{F}_{2021}(0.287)$ is higher than in previous years ( $\mathrm{F}_{2020}=0.239, \mathrm{~F}_{2019}=0.257$ ), and landings have been increasing continuously since 2015, likely because of an increase in recent TACs. Therefore, the working group decided to use the most recent F ( $\mathrm{F}_{2021}=0.287$ ) for the intermediate year (2022). Using a three-year average $F$ would have resulted in a decrease in $F$, which is considered unlikely considering the recent trends in landings and F .

Weights-at-age were calculated as the average of the last three historical years, as in previous years.
The Stock Annex recommends using a long-term geometric mean of the full time-series (1969 to terminal assessment year) for the recruitment in the short-term forecast. However, the Stock Annex also specifies that a short-term geometric mean should be used if distinct periods of successive low or high recruitment are evident over the final three years. In the previous year (WGCSE 2021), such a shorter period was used. However, this year (WGCSE 2022), WGCSE reverted to using the full time-series (the standard practice of the Stock Annex) because recruitment has been highly variably in the recent past with very high $(2018,2020)$ but also very low $(2021)$ recruitment estimates.

The forecast was conducted with FLR's FLash R package using the output from the landings only XSA assessment. The resulting yield was obtained by adding discards to the landing with an average discard rate of the last three historical years (2019-2021, 0.44\%).

The input data for the short-term forecast are shown in Table 32.11.

## MSY forecast

Table 32.12 shows a detailed output of the forecast targeting Fmsy for 2023-2024, and Table 32.13 the year classes contributing to the forecast yield and SSB.

Figure 32.25 shows the forecast results for Fmsy, and Figure 32.26 the forecast, including Fmsy ranges.

Implementing the MSY approach with $\mathrm{Fms} \mathrm{\gamma}_{\mathrm{m}}=0.29$ leads to a total yield of 1394 t in 2023, and an SSB of 4541 t in 2024.

The advice is a reduction of $23 \%$ compared to the advice for 2022. This decline is because of a projected decline in the SSB, caused by (1) low recruitment in 2021 and (2) previously strong cohorts which had increased the SSB are now disappearing because they have been fished down in recent years with the increasing fishing mortality and due to natural mortality reducing the cohorts. This trend is illustrated in Figures 32.27 (biomass) and 32.27 (numbers). The recent increases in SSB have been supported by strong year classes of relatively old and heavier fish, which are disappearing and causing a reduction in the SSB in the short-term forecast (Figure 32.27). This effect is more pronounced in terms of the numbers of fish, which are already declining in 2021 and this trend continues into the short-term forecast (Figure 32.28).

## Additional options

A management options table is provided in Table 32.14, and Table 32.15 shows additional options.

### 32.7 Biological reference points

The most recent reference points for this stock were developed by WKMSYREF4 in 2015 (ICES, 2016). These reference points are presented in the following table. Please note that ICES changed the basis for $\mathrm{F}_{\mathrm{pa}}$ to $\mathrm{F}_{\mathrm{p} .05}$ in 2021, and the updated $\mathrm{F}_{\mathrm{pa}}$ value is shown here.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 2900 t | The 5th percentile of the distribution of SSB when fishing at $\mathrm{F}_{\text {MSY }}$ (0.29) with no error. | ICES $(2016,2017)$ |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.29 | The peak of the median landings yield curve. | ICES $(2016,2017)$ |
|  | $\mathrm{F}_{\text {MSY }}$ lower | 0.16 | Minimum F which produces at least 95\% of maximum yield. | ICES $(2016,2017)$ |
|  | $\mathrm{F}_{\text {MSY }}$ upper | 0.34 | Maximum F which produces at least 95\% of maximum yield. | ICES $(2016,2017)$ |
| Precautionary approach | $\mathrm{Blim}_{\text {lim }}$ | 2000 t | Rounded $\mathrm{B}_{\mathrm{pa}} / 1.4$. | ICES $(2016,2017)$ |
|  | $B_{p a}$ | 2900 t | Rounded $\mathrm{B}_{\text {loss }}$ (1999 year class). Lowest SSB with high recruitment. | ICES $(2016,2017)$ |
|  | $\mathrm{F}_{\text {lim }}$ | 0.44 | Segmented regression simulation of recruitment with $\mathrm{B}_{\mathrm{lim}}$ as the breakpoint and no error. | ICES $(2016,2017)$ |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.39 | Fp.05; the F that leads to SSB $\geq$ Blim with $95 \%$ probability. | ICES $(2016,2017)$ |
| Previous management plan | $\mathrm{SSB}_{\text {MGT }}$ | Not defined |  |  |
|  | $\mathrm{F}_{\text {MGT }}$ | 0.27 |  | EU (2007) |

### 32.8 Management plan

The European Commission implemented a management plan for the recovery of the stock early in 2007 (Council Regulation (EC) No 509/2007). The management plan has not been formally evaluated, but the working group concluded that: The long-term management target (Fмgт = 0.27 ) is precautionary in the sense that it ensures that there is a less than $5 \%$ chance of SSB declining below previously observed levels, as well as maintaining yield within $10 \%$ of MSY (WGCSE note: long-term yield at $F_{M A X}$ (working group, 2005; working group, 2006).

This management plan has not been used in recent years, and the ICES advice has been based on the MSY approach, targeting Fmš.

The management plan (Council Regulation (EC) No 509/2007) is no longer in force since 2019 and has been repealed by an EU multiannual plan for stocks fished in the Western Waters and adjacent waters (Regulation (EU) 2019/472, EU, 2019) which aims at targeting MSY.

### 32.9 Uncertainties in assessment and forecast

The methodology provided is as robust as possible and does not currently appear to suffer from a serious retrospective pattern.

## Discarding

Discarding is considered negligible in the sole fishery, averaging only $0.44 \%$ of the total international catch weight in 2021. Nevertheless, a time-series of available discards information raised to the fleet level should be developed to effectively deal with potential future discard issues and improve estimates of total mortality. The EU landing obligation was implemented during 20162019 with a discard plan and seemed to have reduced the already low discards even more. The landings advice has been topped up with the available discard information to give catch advice. Developing a time-series of discard information appears to be less urgent than in the past.

## Surveys

The assessment methodology includes two survey indices. The UK-Q1SWBeam survey added to the assessment in 2012 covers the entire management area, providing fishery-independent tuning information for the entire age range used in the assessment. Therefore, the assessment relies much less on the commercial tuning information and is less susceptible to localised exploitation by the fishery. Consequently, commercial tuning information is still used in the assessment to maintain the balance between accuracy and precision required by management. Survey information for the recruiting year class remains temporally variable and is not used in the forecast for this reason.

## Sampling

Age and length sampling for this stock is mostly adequate. Age data from the largest sector operating in this fishery (UK) are included in the assessment. France submitted no age samples for the 2019 season. Due to reprocessing of age data submitted by France in 2018, age samples from several strata were retracted. French age data between 2009 and 2014 were insufficient at older ages to raise the length compositions, and therefore UK age data were used to cover the larger fish.

There are very limited discard age samples, but this does not impose a problem on the assessment or forecast due to very low discarding.

## Consistency

The assessment for this stock was last benchmarked in 2012, and an inter-benchmark was held in 2015. The 2022 assessment is consistent with the previous assessments conducted in recent years. Temporal trends in SSB and F estimates were virtually identical.

### 32.10 Recommendation for the next benchmark

There is no requirement to benchmark this stock in the short term.
The XSA assessment uses a taper range of 15 years for the tuning indices, effectively downweighting older tuning data and removing data older than 15 years altogether. As tuning time-
series become longer, potentially important information might get lost in the process. Therefore, a re-evaluation of assessment parametrisation should be considered.

LPUE estimates for the UK-CBT and UK-COT fleets should be closely monitored to avoid the recurrence of inaccuracies in commercial tuning information observed at the 2014 and 2015 working groups. A rescaling observed in the 2018 and 2019 assessments can be explained by underlying data. Consequently, the next benchmark should evaluate the temporal stability of the retrospective patterns and determine whether the assessment settings need to be revised.
The UK-COT effort has been in continuous decline and reported no activity in 2016 and subsequently, due to a new database system, cannot be replicated anymore. Consequently, a benchmark could investigate the removal of commercial tuning information altogether from the assessment.

As the time-series on discards increases, a future benchmark might look into including discard estimates in the assessment and estimating historical discards. Discards are very low and, due to the implementation of the landing obligation in 2016, unlikely to become a problem in the future.

### 32.11 Management considerations

France provided discard estimates for the first time at the 2016 working group. Discard estimates from France are higher than from the other countries.

Plaice is taken as bycatch in this fishery, and therefore management advice for sole must also consider the advice for plaice. Anglerfish, cuttlefish, and lemon sole are also important bycatches in this fishery.

### 32.12 Ecosystem considerations and changes in the environment

See Stock Annex.

### 32.13 Regulations and their effects

Management of this stock is mainly by TAC. In 2005, effort restrictions were implemented for beam trawlers and entangling gears targeting sole in this fishery to enforce the TAC and improve data quality. The effort restrictions were included in the 2007 management plan (EU, 2007) and are continued in the EU multiannual plan (EU, 2019). The effort restrictions limit the number of days at sea for vessels in 7.e using beam trawls ( $\geq 80 \mathrm{~mm}$ mesh size) and static nets ( $\leq 120 \mathrm{~mm}$ mesh size). The limits for effort are set annually in the EU council with the TAC and apply only for vessels which catch more than 300 kg of sole annually.

Mesh restrictions for towed gears are set to 80 mm codends, which correspond well with the minimum landing size of sole at 24 cm ( 25 cm for Belgian vessels since December 2017).

### 32.14 References

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### 32.15 Tables

Table 32.1. Sole in Division 7.e. History of official landings and ICES estimates. All weights are in tonnes.
$\left.\begin{array}{lllllll}\hline \text { Year } & \text { Belgium } & \text { France } & \text { Netherlands } & \text { Ireland } & \begin{array}{l}\text { UK and Chan- } \\ \text { nel Islands }\end{array} & \begin{array}{l}\text { Official to- } \\ \text { tal }\end{array} \\ \hline \text { ICES lands } \\ \text { ings }\end{array}\right]$
$\left.\begin{array}{llllllll}\hline \text { Year } & \text { Belgium } & \text { France } & \text { Netherlands } & \text { Ireland } & \begin{array}{l}\text { UK and Chan- } \\ \text { nel Islands }\end{array} & \begin{array}{l}\text { Official to- } \\ \text { tal }\end{array} & \begin{array}{l}\text { ICES land- } \\ \text { ings }\end{array} \\ \hline 2000 & 4 & 241 & & 413 & 658 & 914 & \\ \hline 2001 & 19 & 224 & & 407 & 650 & 1069 \\ \hline \text { cards dis- }\end{array}\right]$

[^17]Table 32.2. Sole in Division 7.e. Landings numbers-at-age (thousands).

| YEAR\AGE | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 89 | 322 | 80 | 148 | 210 | 21 | 50 | 26 | 20 | 9 | 63 | 1037 |
| 1970 | 53 | 232 | 322 | 90 | 83 | 112 | 13 | 35 | 52 | 22 | 113 | 1127 |
| 1971 | 51 | 200 | 246 | 198 | 65 | 80 | 156 | 10 | 35 | 54 | 113 | 1207 |
| 1972 | 146 | 412 | 167 | 115 | 112 | 14 | 25 | 134 | 38 | 54 | 106 | 1323 |
| 1973 | 71 | 396 | 433 | 89 | 99 | 120 | 17 | 52 | 30 | 4 | 136 | 1446 |
| 1974 | 45 | 349 | 220 | 178 | 71 | 80 | 43 | 32 | 24 | 55 | 106 | 1202 |
| 1975 | 82 | 567 | 170 | 199 | 115 | 28 | 53 | 26 | 22 | 24 | 171 | 1456 |
| 1976 | 167 | 419 | 472 | 161 | 135 | 92 | 46 | 58 | 51 | 14 | 213 | 1830 |
| 1977 | 426 | 318 | 384 | 206 | 102 | 70 | 74 | 10 | 24 | 32 | 159 | 1804 |
| 1978 | 250 | 1123 | 347 | 214 | 189 | 103 | 72 | 77 | 38 | 27 | 203 | 2644 |
| 1979 | 227 | 803 | 811 | 250 | 229 | 174 | 103 | 90 | 104 | 28 | 290 | 3108 |
| 1980 | 175 | 559 | 497 | 630 | 126 | 183 | 140 | 65 | 56 | 130 | 342 | 2902 |
| 1981 | 245 | 806 | 651 | 467 | 389 | 179 | 126 | 76 | 58 | 55 | 211 | 3262 |
| 1982 | 128 | 1451 | 916 | 553 | 352 | 240 | 136 | 113 | 81 | 61 | 294 | 4324 |
| 1983 | 91 | 753 | 1573 | 583 | 351 | 267 | 294 | 119 | 73 | 37 | 262 | 4401 |
| 1984 | 333 | 663 | 826 | 758 | 325 | 204 | 129 | 152 | 54 | 28 | 255 | 3727 |
| 1985 | 287 | 1700 | 756 | 469 | 585 | 179 | 97 | 103 | 85 | 29 | 125 | 4414 |
| 1986 | 246 | 1618 | 971 | 421 | 321 | 336 | 84 | 75 | 90 | 74 | 127 | 4363 |
| 1987 | 487 | 808 | 1090 | 427 | 204 | 224 | 229 | 47 | 50 | 41 | 162 | 3770 |
| 1988 | 443 | 1438 | 596 | 728 | 374 | 153 | 162 | 109 | 39 | 50 | 171 | 4262 |
| 1989 | 390 | 871 | 1233 | 497 | 509 | 225 | 110 | 107 | 113 | 48 | 214 | 4316 |
| 1990 | 341 | 902 | 581 | 553 | 244 | 264 | 143 | 103 | 75 | 85 | 235 | 3525 |
| 1991 | 450 | 415 | 482 | 289 | 220 | 93 | 111 | 68 | 37 | 31 | 145 | 2341 |
| 1992 | 316 | 1434 | 417 | 297 | 115 | 112 | 61 | 74 | 26 | 23 | 90 | 2964 |
| 1993 | 209 | 704 | 1107 | 350 | 219 | 151 | 78 | 60 | 56 | 31 | 79 | 3045 |
| 1994 | 97 | 657 | 558 | 558 | 112 | 106 | 49 | 57 | 44 | 50 | 99 | 2388 |
| 1995 | 95 | 308 | 629 | 427 | 411 | 131 | 101 | 61 | 33 | 18 | 142 | 2356 |
| 1996 | 365 | 445 | 364 | 298 | 235 | 257 | 68 | 61 | 49 | 37 | 143 | 2321 |
| 1997 | 216 | 831 | 724 | 325 | 180 | 194 | 173 | 44 | 20 | 40 | 88 | 2835 |


| YEAR\AGE | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 265 | 606 | 536 | 336 | 209 | 151 | 80 | 127 | 35 | 34 | 162 | 2543 |
| 1999 | 280 | 915 | 500 | 398 | 255 | 114 | 103 | 54 | 107 | 25 | 123 | 2874 |
| 2000 | 307 | 599 | 751 | 367 | 229 | 107 | 53 | 68 | 51 | 88 | 91 | 2710 |
| 2001 | 145 | 1401 | 531 | 497 | 268 | 178 | 100 | 55 | 43 | 42 | 159 | 3419 |
| 2002 | 332 | 1251 | 843 | 387 | 322 | 129 | 105 | 94 | 33 | 18 | 85 | 3599 |
| 2003 | 598 | 835 | 953 | 645 | 130 | 74 | 50 | 58 | 63 | 14 | 61 | 3482 |
| 2004 | 398 | 1080 | 448 | 445 | 526 | 164 | 116 | 61 | 54 | 35 | 85 | 3412 |
| 2005 | 258 | 468 | 834 | 449 | 366 | 293 | 113 | 80 | 45 | 24 | 96 | 3027 |
| 2006 | 500 | 786 | 472 | 606 | 250 | 224 | 185 | 85 | 56 | 31 | 87 | 3282 |
| 2007 | 201 | 852 | 755 | 293 | 362 | 179 | 130 | 110 | 55 | 27 | 99 | 3062 |
| 2008 | 281 | 752 | 678 | 376 | 163 | 184 | 105 | 71 | 67 | 39 | 89 | 2805 |
| 2009 | 166 | 540 | 385 | 333 | 202 | 66 | 74 | 37 | 50 | 35 | 65 | 1955 |
| 2010 | 68 | 348 | 394 | 329 | 204 | 127 | 49 | 71 | 20 | 34 | 78 | 1723 |
| 2011 | 91 | 499 | 476 | 405 | 233 | 156 | 80 | 39 | 34 | 28 | 93 | 2136 |
| 2012 | 31 | 227 | 525 | 400 | 355 | 231 | 137 | 67 | 44 | 39 | 124 | 2180 |
| 2013 | 120 | 324 | 483 | 595 | 280 | 214 | 147 | 98 | 48 | 23 | 110 | 2441 |
| 2014 | 198 | 320 | 466 | 426 | 410 | 168 | 112 | 79 | 61 | 27 | 97 | 2364 |
| 2015 | 177 | 329 | 395 | 336 | 261 | 206 | 115 | 78 | 45 | 30 | 82 | 2054 |
| 2016 | 92 | 420 | 469 | 276 | 249 | 242 | 189 | 67 | 50 | 33 | 107 | 2194 |
| 2017 | 123 | 1188 | 334 | 307 | 277 | 130 | 94 | 41 | 36 | 129 | 78 | 2737 |
| 2018 | 80 | 446 | 410 | 272 | 339 | 156 | 242 | 99 | 82 | 221 | 154 | 2501 |
| 2019 | 115 | 874 | 659 | 633 | 381 | 198 | 168 | 123 | 70 | 86 | 157 | 3463 |
| 2020 | 169 | 558 | 1194 | 613 | 444 | 254 | 137 | 83 | 72 | 33 | 159 | 3716 |
| 2021 | 115 | 841 | 693 | 1182 | 576 | 329 | 173 | 122 | 80 | 48 | 178 | 4338 |

Table 32.3. Sole in Division 7.e. Landings weights-at-age (kg).

| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 0.188 | 0.245 | 0.332 | 0.329 | 0.367 | 0.522 | 0.455 | 0.463 | 0.606 | 0.648 | 0.661 |
| 1970 | 0.188 | 0.224 | 0.295 | 0.315 | 0.355 | 0.436 | 0.500 | 0.444 | 0.514 | 0.530 | 0.596 |
| 1971 | 0.151 | 0.222 | 0.296 | 0.367 | 0.350 | 0.359 | 0.431 | 0.455 | 0.476 | 0.388 | 0.654 |
| 1972 | 0.194 | 0.227 | 0.272 | 0.369 | 0.408 | 0.458 | 0.496 | 0.402 | 0.454 | 0.509 | 0.601 |
| 1973 | 0.203 | 0.224 | 0.262 | 0.311 | 0.382 | 0.415 | 0.460 | 0.467 | 0.538 | 0.655 | 0.562 |
| 1974 | 0.183 | 0.224 | 0.281 | 0.379 | 0.434 | 0.372 | 0.465 | 0.476 | 0.488 | 0.475 | 0.732 |
| 1975 | 0.178 | 0.210 | 0.293 | 0.351 | 0.395 | 0.427 | 0.487 | 0.580 | 0.638 | 0.525 | 0.663 |
| 1976 | 0.170 | 0.218 | 0.287 | 0.324 | 0.391 | 0.455 | 0.414 | 0.476 | 0.479 | 0.585 | 0.629 |
| 1977 | 0.197 | 0.249 | 0.303 | 0.357 | 0.400 | 0.503 | 0.464 | 0.518 | 0.485 | 0.553 | 0.683 |
| 1978 | 0.178 | 0.239 | 0.300 | 0.387 | 0.435 | 0.374 | 0.482 | 0.485 | 0.484 | 0.535 | 0.665 |
| 1979 | 0.189 | 0.239 | 0.330 | 0.427 | 0.464 | 0.472 | 0.481 | 0.570 | 0.527 | 0.574 | 0.732 |
| 1980 | 0.189 | 0.254 | 0.343 | 0.389 | 0.525 | 0.560 | 0.609 | 0.646 | 0.655 | 0.600 | 0.783 |
| 1981 | 0.174 | 0.225 | 0.321 | 0.381 | 0.477 | 0.514 | 0.533 | 0.598 | 0.619 | 0.708 | 0.660 |
| 1982 | 0.214 | 0.209 | 0.278 | 0.347 | 0.426 | 0.498 | 0.510 | 0.523 | 0.526 | 0.564 | 0.663 |
| 1983 | 0.187 | 0.250 | 0.271 | 0.306 | 0.388 | 0.417 | 0.473 | 0.530 | 0.608 | 0.551 | 0.665 |
| 1984 | 0.210 | 0.243 | 0.306 | 0.381 | 0.391 | 0.481 | 0.542 | 0.562 | 0.604 | 0.726 | 0.643 |
| 1985 | 0.163 | 0.226 | 0.298 | 0.360 | 0.391 | 0.472 | 0.523 | 0.534 | 0.522 | 0.588 | 0.822 |
| 1986 | 0.174 | 0.237 | 0.297 | 0.354 | 0.407 | 0.456 | 0.502 | 0.544 | 0.583 | 0.618 | 0.703 |
| 1987 | 0.174 | 0.245 | 0.310 | 0.370 | 0.425 | 0.474 | 0.518 | 0.557 | 0.590 | 0.618 | 0.665 |
| 1988 | 0.170 | 0.244 | 0.312 | 0.375 | 0.432 | 0.484 | 0.531 | 0.572 | 0.608 | 0.639 | 0.694 |
| 1989 | 0.167 | 0.222 | 0.275 | 0.326 | 0.375 | 0.422 | 0.467 | 0.510 | 0.551 | 0.590 | 0.692 |
| 1990 | 0.217 | 0.272 | 0.324 | 0.372 | 0.419 | 0.461 | 0.501 | 0.538 | 0.571 | 0.601 | 0.669 |
| 1991 | 0.182 | 0.255 | 0.323 | 0.386 | 0.445 | 0.499 | 0.549 | 0.594 | 0.634 | 0.669 | 0.741 |
| 1992 | 0.166 | 0.238 | 0.305 | 0.366 | 0.423 | 0.474 | 0.520 | 0.561 | 0.597 | 0.627 | 0.683 |
| 1993 | 0.146 | 0.209 | 0.268 | 0.324 | 0.376 | 0.425 | 0.470 | 0.513 | 0.551 | 0.587 | 0.672 |
| 1994 | 0.183 | 0.241 | 0.295 | 0.347 | 0.396 | 0.442 | 0.484 | 0.524 | 0.561 | 0.595 | 0.671 |
| 1995 | 0.192 | 0.248 | 0.301 | 0.351 | 0.397 | 0.441 | 0.481 | 0.518 | 0.552 | 0.583 | 0.652 |
| 1996 | 0.214 | 0.262 | 0.308 | 0.354 | 0.399 | 0.442 | 0.484 | 0.524 | 0.564 | 0.602 | 0.694 |
| 1997 | 0.186 | 0.244 | 0.300 | 0.354 | 0.406 | 0.455 | 0.503 | 0.548 | 0.592 | 0.633 | 0.734 |


| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 0.191 | 0.247 | 0.300 | 0.350 | 0.397 | 0.441 | 0.482 | 0.520 | 0.555 | 0.586 | 0.661 |
| 1999 | 0.208 | 0.257 | 0.303 | 0.347 | 0.389 | 0.429 | 0.468 | 0.503 | 0.536 | 0.567 | 0.637 |
| 2000 | 0.202 | 0.258 | 0.310 | 0.358 | 0.401 | 0.441 | 0.476 | 0.508 | 0.535 | 0.558 | 0.647 |
| 2001 | 0.203 | 0.245 | 0.287 | 0.326 | 0.365 | 0.402 | 0.438 | 0.472 | 0.505 | 0.537 | 0.616 |
| 2002 | 0.181 | 0.236 | 0.290 | 0.342 | 0.391 | 0.439 | 0.485 | 0.529 | 0.570 | 0.610 | 0.706 |
| 2003 | 0.173 | 0.241 | 0.306 | 0.367 | 0.425 | 0.479 | 0.530 | 0.577 | 0.620 | 0.660 | 0.746 |
| 2004 | 0.176 | 0.230 | 0.282 | 0.334 | 0.385 | 0.435 | 0.485 | 0.534 | 0.582 | 0.629 | 0.757 |
| 2005 | 0.180 | 0.236 | 0.290 | 0.343 | 0.394 | 0.444 | 0.493 | 0.540 | 0.586 | 0.630 | 0.747 |
| 2006 | 0.169 | 0.228 | 0.282 | 0.333 | 0.381 | 0.424 | 0.464 | 0.501 | 0.533 | 0.562 | 0.672 |
| 2007 | 0.183 | 0.244 | 0.299 | 0.350 | 0.395 | 0.436 | 0.471 | 0.501 | 0.526 | 0.546 | 0.616 |
| 2008 | 0.197 | 0.245 | 0.292 | 0.337 | 0.382 | 0.425 | 0.468 | 0.509 | 0.549 | 0.588 | 0.652 |
| 2009 | 0.176 | 0.252 | 0.322 | 0.385 | 0.443 | 0.494 | 0.540 | 0.579 | 0.612 | 0.639 | 0.703 |
| 2010 | 0.169 | 0.258 | 0.339 | 0.412 | 0.476 | 0.532 | 0.580 | 0.619 | 0.650 | 0.673 | 0.699 |
| 2011 | 0.200 | 0.261 | 0.319 | 0.375 | 0.428 | 0.480 | 0.528 | 0.575 | 0.618 | 0.660 | 0.749 |
| 2012 | 0.162 | 0.240 | 0.311 | 0.373 | 0.428 | 0.476 | 0.516 | 0.548 | 0.572 | 0.589 | 0.664 |
| 2013 | 0.172 | 0.228 | 0.283 | 0.337 | 0.389 | 0.439 | 0.489 | 0.536 | 0.583 | 0.628 | 0.740 |
| 2014 | 0.191 | 0.254 | 0.313 | 0.366 | 0.415 | 0.459 | 0.499 | 0.533 | 0.563 | 0.588 | 0.709 |
| 2015 | 0.182 | 0.250 | 0.313 | 0.370 | 0.423 | 0.471 | 0.513 | 0.551 | 0.583 | 0.611 | 0.697 |
| 2016 | 0.215 | 0.282 | 0.345 | 0.401 | 0.453 | 0.499 | 0.541 | 0.576 | 0.606 | 0.631 | 0.720 |
| 2017 | 0.225 | 0.279 | 0.331 | 0.382 | 0.432 | 0.479 | 0.525 | 0.568 | 0.610 | 0.651 | 0.763 |
| 2018 | 0.205 | 0.264 | 0.321 | 0.374 | 0.425 | 0.473 | 0.518 | 0.560 | 0.600 | 0.636 | 0.768 |
| 2019 | 0.180 | 0.233 | 0.284 | 0.333 | 0.379 | 0.423 | 0.464 | 0.503 | 0.540 | 0.574 | 0.682 |
| 2020 | 0.188 | 0.235 | 0.280 | 0.323 | 0.365 | 0.406 | 0.445 | 0.483 | 0.519 | 0.553 | 0.642 |
| 2021 | 0.157 | 0.213 | 0.265 | 0.313 | 0.358 | 0.399 | 0.435 | 0.468 | 0.498 | 0.523 | 0.592 |

Table 32.4. Sole in Division 7.e. Stock weights-at-age (kg).

| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 0.125 | 0.200 | 0.270 | 0.330 | 0.380 | 0.425 | 0.460 | 0.490 | 0.520 | 0.550 | 0.609 |
| 1970 | 0.120 | 0.195 | 0.255 | 0.305 | 0.355 | 0.395 | 0.430 | 0.465 | 0.490 | 0.510 | 0.541 |
| 1971 | 0.090 | 0.170 | 0.240 | 0.295 | 0.345 | 0.390 | 0.420 | 0.445 | 0.470 | 0.490 | 0.544 |
| 1972 | 0.130 | 0.200 | 0.265 | 0.325 | 0.380 | 0.420 | 0.460 | 0.490 | 0.520 | 0.540 | 0.558 |
| 1973 | 0.105 | 0.170 | 0.235 | 0.290 | 0.340 | 0.390 | 0.435 | 0.475 | 0.510 | 0.540 | 0.585 |
| 1974 | 0.125 | 0.200 | 0.265 | 0.320 | 0.370 | 0.410 | 0.455 | 0.490 | 0.515 | 0.530 | 0.571 |
| 1975 | 0.144 | 0.221 | 0.267 | 0.327 | 0.385 | 0.435 | 0.479 | 0.516 | 0.545 | 0.569 | 0.628 |
| 1976 | 0.146 | 0.198 | 0.247 | 0.294 | 0.338 | 0.380 | 0.417 | 0.456 | 0.491 | 0.523 | 0.595 |
| 1977 | 0.156 | 0.221 | 0.278 | 0.332 | 0.382 | 0.425 | 0.462 | 0.497 | 0.527 | 0.553 | 0.629 |
| 1978 | 0.156 | 0.217 | 0.276 | 0.330 | 0.380 | 0.425 | 0.463 | 0.498 | 0.526 | 0.555 | 0.630 |
| 1979 | 0.141 | 0.216 | 0.287 | 0.352 | 0.414 | 0.463 | 0.502 | 0.539 | 0.574 | 0.608 | 0.719 |
| 1980 | 0.125 | 0.206 | 0.288 | 0.360 | 0.436 | 0.513 | 0.575 | 0.620 | 0.650 | 0.674 | 0.714 |
| 1981 | 0.119 | 0.197 | 0.276 | 0.358 | 0.427 | 0.490 | 0.543 | 0.582 | 0.616 | 0.645 | 0.699 |
| 1982 | 0.117 | 0.195 | 0.265 | 0.335 | 0.398 | 0.455 | 0.506 | 0.536 | 0.562 | 0.585 | 0.632 |
| 1983 | 0.120 | 0.195 | 0.250 | 0.307 | 0.365 | 0.420 | 0.475 | 0.520 | 0.570 | 0.615 | 0.709 |
| 1984 | 0.108 | 0.192 | 0.268 | 0.339 | 0.400 | 0.453 | 0.501 | 0.545 | 0.577 | 0.607 | 0.696 |
| 1985 | 0.150 | 0.204 | 0.258 | 0.311 | 0.364 | 0.416 | 0.468 | 0.520 | 0.571 | 0.621 | 0.790 |
| 1986 | 0.140 | 0.206 | 0.268 | 0.326 | 0.381 | 0.432 | 0.480 | 0.524 | 0.564 | 0.601 | 0.691 |
| 1987 | 0.137 | 0.210 | 0.278 | 0.341 | 0.398 | 0.450 | 0.497 | 0.538 | 0.574 | 0.605 | 0.659 |
| 1988 | 0.131 | 0.208 | 0.278 | 0.344 | 0.404 | 0.459 | 0.508 | 0.552 | 0.591 | 0.624 | 0.687 |
| 1989 | 0.139 | 0.195 | 0.249 | 0.300 | 0.350 | 0.398 | 0.444 | 0.488 | 0.531 | 0.571 | 0.675 |
| 1990 | 0.187 | 0.243 | 0.296 | 0.346 | 0.393 | 0.437 | 0.478 | 0.516 | 0.551 | 0.583 | 0.654 |
| 1991 | 0.144 | 0.219 | 0.290 | 0.355 | 0.416 | 0.473 | 0.524 | 0.572 | 0.614 | 0.652 | 0.731 |
| 1992 | 0.128 | 0.202 | 0.272 | 0.336 | 0.395 | 0.449 | 0.498 | 0.542 | 0.580 | 0.613 | 0.677 |
| 1993 | 0.114 | 0.178 | 0.239 | 0.296 | 0.350 | 0.401 | 0.448 | 0.492 | 0.532 | 0.570 | 0.659 |
| 1994 | 0.153 | 0.212 | 0.268 | 0.322 | 0.372 | 0.419 | 0.463 | 0.505 | 0.543 | 0.578 | 0.659 |
| 1995 | 0.163 | 0.221 | 0.275 | 0.326 | 0.374 | 0.419 | 0.461 | 0.500 | 0.536 | 0.568 | 0.641 |
| 1996 | 0.189 | 0.238 | 0.285 | 0.331 | 0.376 | 0.420 | 0.463 | 0.504 | 0.544 | 0.583 | 0.677 |
| 1997 | 0.156 | 0.215 | 0.272 | 0.327 | 0.380 | 0.431 | 0.480 | 0.526 | 0.570 | 0.612 | 0.717 |


| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 0.162 | 0.220 | 0.274 | 0.325 | 0.374 | 0.419 | 0.462 | 0.501 | 0.537 | 0.571 | 0.650 |
| 1999 | 0.183 | 0.233 | 0.280 | 0.326 | 0.369 | 0.410 | 0.448 | 0.485 | 0.519 | 0.551 | 0.624 |
| 2000 | 0.172 | 0.230 | 0.284 | 0.333 | 0.379 | 0.421 | 0.458 | 0.492 | 0.521 | 0.546 | 0.643 |
| 2001 | 0.181 | 0.224 | 0.266 | 0.307 | 0.346 | 0.384 | 0.420 | 0.455 | 0.489 | 0.521 | 0.602 |
| 2002 | 0.152 | 0.209 | 0.263 | 0.316 | 0.367 | 0.415 | 0.462 | 0.507 | 0.550 | 0.591 | 0.688 |
| 2003 | 0.137 | 0.207 | 0.274 | 0.337 | 0.396 | 0.452 | 0.505 | 0.554 | 0.599 | 0.641 | 0.732 |
| 2004 | 0.149 | 0.203 | 0.256 | 0.308 | 0.360 | 0.410 | 0.460 | 0.509 | 0.557 | 0.605 | 0.734 |
| 2005 | 0.152 | 0.208 | 0.263 | 0.316 | 0.368 | 0.419 | 0.468 | 0.516 | 0.562 | 0.607 | 0.726 |
| 2006 | 0.138 | 0.197 | 0.254 | 0.306 | 0.355 | 0.400 | 0.442 | 0.479 | 0.514 | 0.544 | 0.661 |
| 2007 | 0.151 | 0.214 | 0.272 | 0.325 | 0.373 | 0.416 | 0.454 | 0.486 | 0.514 | 0.536 | 0.614 |
| 2008 | 0.172 | 0.221 | 0.268 | 0.315 | 0.360 | 0.404 | 0.447 | 0.489 | 0.529 | 0.569 | 0.640 |
| 2009 | 0.136 | 0.215 | 0.287 | 0.354 | 0.415 | 0.469 | 0.518 | 0.560 | 0.596 | 0.626 | 0.698 |
| 2010 | 0.121 | 0.215 | 0.300 | 0.376 | 0.445 | 0.505 | 0.557 | 0.600 | 0.636 | 0.663 | 0.696 |
| 2011 | 0.169 | 0.231 | 0.290 | 0.347 | 0.402 | 0.454 | 0.504 | 0.552 | 0.597 | 0.639 | 0.738 |
| 2012 | 0.120 | 0.202 | 0.276 | 0.343 | 0.402 | 0.453 | 0.497 | 0.532 | 0.561 | 0.581 | 0.664 |
| 2013 | 0.144 | 0.200 | 0.256 | 0.310 | 0.363 | 0.414 | 0.464 | 0.513 | 0.560 | 0.606 | 0.729 |
| 2014 | 0.157 | 0.223 | 0.284 | 0.340 | 0.391 | 0.438 | 0.480 | 0.517 | 0.549 | 0.576 | 0.706 |
| 2015 | 0.147 | 0.217 | 0.282 | 0.342 | 0.397 | 0.448 | 0.493 | 0.533 | 0.568 | 0.598 | 0.692 |
| 2016 | 0.178 | 0.248 | 0.313 | 0.373 | 0.427 | 0.476 | 0.519 | 0.557 | 0.590 | 0.617 | 0.714 |
| 2017 | 0.197 | 0.252 | 0.305 | 0.357 | 0.407 | 0.455 | 0.501 | 0.546 | 0.588 | 0.630 | 0.749 |
| 2018 | 0.174 | 0.235 | 0.293 | 0.348 | 0.400 | 0.450 | 0.496 | 0.540 | 0.580 | 0.618 | 0.760 |
| 2019 | 0.152 | 0.207 | 0.259 | 0.309 | 0.356 | 0.401 | 0.444 | 0.484 | 0.522 | 0.557 | 0.672 |
| 2020 | 0.165 | 0.212 | 0.257 | 0.302 | 0.344 | 0.386 | 0.426 | 0.464 | 0.501 | 0.536 | 0.632 |
| 2021 | 0.128 | 0.186 | 0.240 | 0.290 | 0.336 | 0.379 | 0.417 | 0.452 | 0.483 | 0.511 | 0.586 |

Table 32.5. Sole in Division 7.e. Landings, effort and mean standardised Ipue for the UK commercial fleets.

| Fleet | Year | Effort [days] | Landings [tonnes] | Lpue [tonnes/1000 days] | means standardised Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
| UK-CBT<24 m | 1988 | 2527 | 293 | 115.97 | 1.93 |
|  | 1989 | 1956 | 162 | 83.06 | 1.38 |
|  | 1990 | 1958 | 179 | 91.51 | 1.52 |
|  | 1991 | 1458 | 134 | 92.22 | 1.53 |
|  | 1992 | 1342 | 142 | 106.22 | 1.76 |
|  | 1993 | 1432 | 154 | 107.71 | 1.79 |
|  | 1994 | 2241 | 161 | 71.97 | 1.19 |
|  | 1995 | 2017 | 134 | 66.28 | 1.1 |
|  | 1996 | 1999 | 106 | 52.99 | 0.88 |
|  | 1997 | 1991 | 132 | 66.3 | 1.1 |
|  | 1998 | 2357 | 99 | 42.12 | 0.7 |
|  | 1999 | 2518 | 115 | 45.7 | 0.76 |
|  | 2000 | 2913 | 134 | 45.85 | 0.76 |
|  | 2001 | 3746 | 148 | 39.57 | 0.66 |
|  | 2002 | 3482 | 110 | 31.55 | 0.52 |
|  | 2003 | 3785 | 93 | 24.44 | 0.41 |
|  | 2004 | 3512 | 64 | 18.12 | 0.3 |
|  | 2005 | 3305 | 191 | 57.72 | 0.96 |
|  | 2006 | 3277 | 224 | 68.27 | 1.13 |
|  | 2007 | 4027 | 225 | 55.77 | 0.93 |
|  | 2008 | 4629 | 213 | 45.94 | 0.76 |
|  | 2009 | 4040 | 185 | 45.85 | 0.76 |
|  | 2010 | 4727 | 201 | 42.42 | 0.7 |
|  | 2011 | 5913 | 258 | 43.65 | 0.72 |
|  | 2012 | 7188 | 314 | 43.65 | 0.72 |
|  | 2013 | 6322 | 329 | 52.02 | 0.86 |
|  | 2014 | 5870 | 308 | 52.54 | 0.87 |
|  | 2015 | 6260 | 310 | 49.54 | 0.82 |


| Fleet | Year | Effort [days] | Landings [tonnes] | Lpue [tonnes/1000 days] | means standardised Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2016 | 6114 | 355 | 58.1 | 0.96 |
|  | 2017 | 6578 | 402 | 61.08 | 1.01 |
|  | 2018 | 6366 | 386 | 60.66 | 1.01 |
|  | 2019 | 6067 | 397 | 65.49 | 1.09 |
|  | 2020 | 5643 | 393 | 69.61 | 1.16 |
|  | 2021 | 5354 | 399 | 74.5 | 1.24 |
| UK-CBT>24 m | 1988 | 2971 | 391 | 131.77 | 2.74 |
|  | 1989 | 3938 | 340 | 86.37 | 1.79 |
|  | 1990 | 3518 | 314 | 89.12 | 1.85 |
|  | 1991 | 2412 | 206 | 85.47 | 1.78 |
|  | 1992 | 1993 | 197 | 98.63 | 2.05 |
|  | 1993 | 2678 | 194 | 72.54 | 1.51 |
|  | 1994 | 4574 | 236 | 51.5 | 1.07 |
|  | 1995 | 4917 | 257 | 52.3 | 1.09 |
|  | 1996 | 5592 | 178 | 31.84 | 0.66 |
|  | 1997 | 5377 | 199 | 37.1 | 0.77 |
|  | 1998 | 4945 | 164 | 33.19 | 0.69 |
|  | 1999 | 4512 | 141 | 31.32 | 0.65 |
|  | 2000 | 5237 | 151 | 28.84 | 0.6 |
|  | 2001 | 5874 | 142 | 24.11 | 0.5 |
|  | 2002 | 5957 | 104 | 17.51 | 0.36 |
|  | 2003 | 6810 | 94 | 13.78 | 0.29 |
|  | 2004 | 7100 | 69 | 9.66 | 0.2 |
|  | 2005 | 6684 | 236 | 35.27 | 0.73 |
|  | 2006 | 6595 | 236 | 35.79 | 0.74 |
|  | 2007 | 5594 | 196 | 35.1 | 0.73 |
|  | 2008 | 4924 | 154 | 31.36 | 0.65 |
|  | 2009 | 3523 | 115 | 32.66 | 0.68 |
|  | 2010 | 3064 | 94 | 30.64 | 0.64 |


| Fleet | Year | Effort [days] | Landings [tonnes] | Lpue [tonnes/1000 days] | means standardised Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2011 | 2790 | 92 | 32.95 | 0.68 |
|  | 2012 | 2609 | 86 | 33.01 | 0.69 |
|  | 2013 | 2444 | 93 | 38.13 | 0.79 |
|  | 2014 | 2900 | 104 | 35.95 | 0.75 |
|  | 2015 | 3039 | 101 | 33.12 | 0.69 |
|  | 2016 | 4064 | 166 | 40.79 | 0.85 |
|  | 2017 | 4556 | 207 | 45.41 | 0.94 |
|  | 2018 | 4116 | 231 | 56.17 | 1.17 |
|  | 2019 | 4329 | 313 | 72.36 | 1.5 |
|  | 2020 | 4335 | 321 | 74.07 | 1.54 |
|  | 2021 | 4505 | 354 | 78.56 | 1.63 |
| UK-CBT | 1988 | 5497 | 684 | 124.51 | 2.34 |
|  | 1989 | 5894 | 503 | 85.27 | 1.6 |
|  | 1990 | 5476 | 493 | 89.97 | 1.69 |
|  | 1991 | 3870 | 341 | 88.02 | 1.65 |
|  | 1992 | 3334 | 339 | 101.69 | 1.91 |
|  | 1993 | 4111 | 349 | 84.79 | 1.59 |
|  | 1994 | 6814 | 397 | 58.23 | 1.09 |
|  | 1995 | 6935 | 391 | 56.37 | 1.06 |
|  | 1996 | 7591 | 284 | 37.41 | 0.7 |
|  | 1997 | 7368 | 331 | 44.99 | 0.84 |
|  | 1998 | 7302 | 263 | 36.07 | 0.68 |
|  | 1999 | 7031 | 256 | 36.47 | 0.68 |
|  | 2000 | 8150 | 285 | 34.92 | 0.66 |
|  | 2001 | 9620 | 290 | 30.13 | 0.57 |
|  | 2002 | 9439 | 214 | 22.69 | 0.43 |
|  | 2003 | 10596 | 186 | 17.59 | 0.33 |
|  | 2004 | 10612 | 132 | 12.46 | 0.23 |
|  | 2005 | 9990 | 427 | 42.7 | 0.8 |


| Fleet | Year | Effort [days] | Landings [tonnes] | Lpue [tonnes/1000 days] | means standardised Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2006 | 9873 | 460 | 46.57 | 0.87 |
|  | 2007 | 9621 | 421 | 43.75 | 0.82 |
|  | 2008 | 9552 | 367 | 38.42 | 0.72 |
|  | 2009 | 7563 | 300 | 39.7 | 0.75 |
|  | 2010 | 7791 | 294 | 37.79 | 0.71 |
|  | 2011 | 8703 | 350 | 40.22 | 0.75 |
|  | 2012 | 9797 | 400 | 40.82 | 0.77 |
|  | 2013 | 8767 | 422 | 48.15 | 0.9 |
|  | 2014 | 8769 | 413 | 47.05 | 0.88 |
|  | 2015 | 9298 | 411 | 44.17 | 0.83 |
|  | 2016 | 10178 | 521 | 51.19 | 0.96 |
|  | 2017 | 11114 | 606 | 54.57 | 1.02 |
|  | 2018 | 10482 | 617 | 58.9 | 1.11 |
|  | 2019 | 10396 | 711 | 68.35 | 1.28 |
|  | 2020 | 9978 | 714 | 71.55 | 1.34 |
|  | 2021 | 9859 | 753 | 76.35 | 1.43 |
| UK-COT | 1988 | 4265 | 29 | 6.77 | 1.43 |
|  | 1989 | 4607 | 28 | 6.18 | 1.31 |
|  | 1990 | 4423 | 26 | 5.97 | 1.27 |
|  | 1991 | 4004 | 14 | 3.39 | 0.72 |
|  | 1992 | 4108 | 12 | 3.02 | 0.64 |
|  | 1993 | 3761 | 15 | 3.95 | 0.84 |
|  | 1994 | 3423 | 18 | 5.27 | 1.12 |
|  | 1995 | 3294 | 13 | 3.99 | 0.84 |
|  | 1996 | 2589 | 12 | 4.83 | 1.02 |
|  | 1997 | 3011 | 15 | 4.96 | 1.05 |
|  | 1998 | 2699 | 11 | 4.22 | 0.89 |
|  | 1999 | 2486 | 13 | 5.16 | 1.09 |
|  | 2000 | 2681 | 11 | 4.11 | 0.87 |


| Fleet | Year | Effort [days] | Landings [tonnes] | Lpue [tonnes/1000 days] | means standardised Ipue |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2001 | 2732 | 13 | 4.9 | 1.04 |
|  | 2002 | 2448 | 9 | 3.66 | 0.78 |
|  | 2003 | 2273 | 8 | 3.31 | 0.7 |
|  | 2004 | 2334 | 6 | 2.46 | 0.52 |
|  | 2005 | 1762 | 12 | 6.86 | 1.45 |
|  | 2006 | 1699 | 8 | 4.57 | 0.97 |
|  | 2007 | 1917 | 9 | 4.9 | 1.04 |
|  | 2008 | 1750 | 7 | 4.26 | 0.9 |
|  | 2009 | 1847 | 10 | 5.36 | 1.14 |
|  | 2010 | 2213 | 10 | 4.53 | 0.96 |
|  | 2011 | 1930 | 8 | 4.08 | 0.86 |
|  | 2012 | 2068 | 12 | 5.96 | 1.26 |
|  | 2013 | 1587 | 8 | 4.96 | 1.05 |
|  | 2014 | 1440 | 8 | 5.56 | 1.18 |
|  | 2015 | 978 | 5 | 4.98 | 1.06 |
|  | 2016 | 0 | 0 | NA | NA |
| UK-COT new | 2016 | 2020 | 14 | 7.08 | 0.71 |
|  | 2017 | 2398 | 15 | 6.1 | 0.61 |
|  | 2018 | 1986 | 17 | 8.42 | 0.84 |
|  | 2019 | 1548 | 14 | 9.33 | 0.93 |
|  | 2020 | 1076 | 15 | 14.23 | 1.43 |
|  | 2021 | 1353 | 20 | 14.75 | 1.48 |

Table 32.6. Sole in Division 7.e. Tuning data file. Not all tuning time-series, years and ages shown here were used in the assessment.

| so1.27.7e WGCSE 2022 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 104 |  |  |  |  |  |  |  |
| UK-CBT-1 ate |  |  |  |  |  |  |  |
| 20032021 |  |  |  |  |  |  |  |
| 1101 |  |  |  |  |  |  |  |
| 314 |  |  |  |  |  |  |  |
| $\begin{array}{ll}10.59557 \\ 6.67 & 2.19\end{array}$ |  | 168.87 | 129.96 | 21.43 | 18.32 | 10.28 | 13.49 |
|  |  | 2.06 | 3.35 | 2.82 |  |  |  |
| 10.61183 | 146.5 | 61.53 | 53.46 | 75.23 | 11.35 | 14.96 | 7.49 |
| 5.98 | 4.27 | 2.12 | 1.18 | 1.89 |  |  |  |
| 9.98951 | 210.39 | 326.3 | 132.94 | 155.21 | 132.09 | 27.41 | 32.6 |
| 22.54 | 14.24 | 8.3 | 5.95 | 4.84 |  |  |  |
| 9.87254 | 376.87 | 186.46 | 243.45 | 85.59 | 108.34 | 106.98 | 37.22 |
| 20.67 | 13.69 | 13.61 | 6.68 | 2.99 |  |  |  |
| 9.6207 | 7456.04 | 261.42 | 105.82 | 103.55 | 54.21 | 62.07 | 51.47 |
| 15.34 | 11.12 | 10.41 | 8.44 | 8.17 |  |  |  |
| 9.55231 | - 294.03 | 286.06 | 126.1 | 67.89 | 65.42 | 42.34 | 39.54 |
| 36.27 | 14.54 | 11.8 | 4.3 | 6 |  |  |  |
| 7.56283 | 190.03 | 182.63 | 152.83 | 89.59 | 26.02 | 27.9 | 13.23 |
| 16.1 | 12.91 | 4.85 | 3.74 | 1.92 |  |  |  |
| 7.79112 | -80.09 | 179.7 | 157.57 | 101.24 | 51.98 | 25.24 | 22.59 |
| 8.23 | 16.75 | 25.39 | 7.42 | 3.88 |  |  |  |
| 8.70287 | 243.76 | 148.58 | 186.66 | 121.43 | 81.66 | 35.56 | 15.79 |
| 20.25 | 10.83 | 14.11 | 8.26 | 2.1 |  |  |  |
| 9.79734 | -129.79 | 307.88 | 139.02 | 143.59 | 91.49 | 66.22 | 30.49 |
| 17.81 | 14.83 | 8.55 | 12.25 | 11.03 |  |  |  |
| 8.76655 | -81.92 | 242.49 | 288.92 | 134.34 | 93.18 | 72.27 | 44.15 |
| 24.5 | 10.73 | 9.84 | 8.14 | 9.84 |  |  |  |
|  | 111.72 | 201.15 | 169.62 | 201.19 | 99.91 | 67.46 | 43.84 |
| 30.63 | 15.94 | 7.71 | 9.34 | 4.9 |  |  |  |
| 9.29849 | -137.05 | 178.21 | 198.83 | 135.74 | 117.19 | 65.74 | 45.95 |
| 31.78 | 20.59 | 11.01 | 5.52 | 5.96 |  |  |  |
| 10.17804 | - 263.46 | 217.34 | 158.93 | 161.88 | 118.88 | 102.14 | 49.07 |
| 45.22 | 21.3 | 23.14 | 13.03 | 5.69 |  |  |  |
| 11.11408 | - 454.27 | 353.27 | 177.37 | 142.06 | 120.28 | 81.72 | 72.95 |
| 42.23 | 28.03 | 16.59 | 11.97 | 9.63 |  |  |  |
| 10.48248 | - 217.63 | 454.82 | 260.75 | 116.59 | 118.4 | 76.79 | 51.54 |
| 49.36 | 33.91 | 24.42 | 21.84 | 10.92 |  |  |  |
| 10.39628 | -618.98 | 411.51 | 357.08 | 217.83 | 105.4 | 69.38 | 57 |
| 36.74 | 40.95 | 22.94 | 13.23 | 10.34 |  |  |  |
| 9.97809 | - 366.92 | 668.85 | 351.1 | 232.9 | 155.35 | 85.53 | 44.65 |
| 28.78 | 13.68 | 12.36 | 13.5 | 10.39 |  |  |  |
| 9.85862 | - 489.94 | 449.21 | 574.6 | 243.75 | 181.18 | 96.72 | 47.65 |
| 40.73 | 26.66 | 25.03 | 24.62 | 14.92 |  |  |  |


| UK-COT |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19882016 |  |  |  |  |  |  |
| 1101 |  |  |  |  |  |  |
| 311 |  |  |  |  |  |  |
| 4264.71 | 30.97 | 15.73 | 19.29 | 8.63 | 2.55 | 2.55 |
|  | 1.83 | 0.35 | 0.76 |  |  |  |
| 4607.04 | 15.09 | 18.34 | 9.22 | 11.75 | 4.72 | 2.42 |
|  | 2.36 | 2.01 | 1.4 |  |  |  |
| 4422.52 | 18.3 | 12.56 | 9.21 | 6.09 | 5.53 | 2.08 |
|  | 1.83 | 1.12 | 0.9 |  |  |  |
| 4004.37 | 10.04 | 7.03 | 4.12 | 2.46 | 0.96 | 1.44 |
|  | 0.42 | 0.41 | 0.23 |  |  |  |
| 4107.71 | 26.24 | 6 | 3.6 | 1.19 | 1.14 | 0.48 |
|  | 0.65 | 0.17 | 0.09 |  |  |  |
| 3761 | 12.45 | 17.56 | 5.38 | 3.44 | 2.49 | 1.26 |
|  | 1 | 0.92 | 0.56 |  |  |  |
| 3423.03 | 12.42 | 11.46 | 12.35 | 2.5 | 2.6 | 1.23 |
|  | 1.35 | 1.03 | 1.18 |  |  |  |
| 3294.06 | 5.25 | 9.75 | 6.34 | 6.17 | 1.89 | 1.49 |
|  | 0.91 | 0.52 | 0.25 |  |  |  |
| 2589.38 | 9.47 | 6.54 | 4.37 | 3.15 | 3.54 | 0.95 |
|  | 0.76 | 0.68 | 0.45 |  |  |  |
| 3010.66 | 15.16 | 8.81 | 4.78 | 2.83 | 2.9 | 2.53 |
|  | 0.63 | 0.28 | 0.43 |  |  |  |
| 2698.6 | 8.74 | 7.58 | 4.25 | 2.49 | 1.53 | 0.93 |
|  | 1.47 | 0.31 | 0.44 |  |  |  |
| 2486.17 | 11.56 | 5.84 | 4.91 | 2.89 | 1.45 | 1.46 |
|  | 0.74 | 1.49 | 0.39 |  |  |  |
| 2680.63 | 6.67 | 8.41 | 4.03 | 2.64 | 1.24 | 0.59 |
|  | 0.81 | 0.62 | 0.99 |  |  |  |
| 2731.54 | 18.02 | 5.27 | 4.96 | 2.69 | 2.01 | 1.12 |
|  | 0.7 | 0.51 | 0.5 |  |  |  |
| 2448.37 | 9.88 | 6.12 | 2.39 | 2.67 | 1.27 | 0.82 |
|  | 0.33 | 0.2 | 0.25 |  |  |  |
| 2272.9 | 4.61 | 5.87 | 4.8 | 1.04 | 0.85 | 0.49 |
|  | 0.54 | 0.27 | 0.13 |  |  |  |
| 2334.16 | 6.05 | 2.58 | 2.23 | 3.25 | 0.46 | 0.57 |
|  | 0.3 | 0.24 | 0.18 |  |  |  |
| 1762.36 | 6.44 | 9.56 | 3.53 | 4.13 | 3.44 | 0.74 |
|  | 0.9 | 0.58 | 0.45 |  |  |  |
| 1699.49 | 6.93 | 3.27 | 4.13 | 1.36 | 1.63 | 1.75 |
|  | 0.6 | 0.31 | 0.2 |  |  |  |
| 1916.84 | 9.32 | 5.44 | 2.3 | 2.32 | 1.19 | 1.41 |
|  | 1.13 | 0.36 | 0.21 |  |  |  |
| 1750.36 | 5.61 | 4.85 | 2.08 | 1.15 | 1.18 | 0.75 |
|  | 0.75 | 0.7 | 0.32 |  |  |  |
| 1847.2 | 7.97 | 5.47 | 3.92 | 2.17 | 0.64 | 0.83 |
|  | 0.39 | 0.52 | 0.45 |  |  |  |
| 2212.85 | 2.71 | 5.85 | 4.74 | 3.15 | 1.63 | 0.81 |
|  | 0.74 | 0.3 | 0.6 |  |  |  |
| 1930.5 | 6.51 | 3.32 | 3.89 | 2.46 | 1.64 | 0.58 |
|  | 0.31 | 0.37 | 0.19 |  |  |  |


| 2068.16 | 4.24 | 9.16 | 3.97 | 4.06 | 2.3 | 1.76 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.82 | 0.49 | 0.46 |  |  |  |
| 1586.58 | 2.01 | 4.55 | 5.64 | 2.66 | 1.74 | 1.49 |
|  | 0.89 | 0.56 | 0.26 |  |  |  |
| 1440.22 | 2.13 | 3.57 | 2.99 | 3.56 | 1.8 | 1.29 |
|  | 0.9 | 0.68 | 0.34 |  |  |  |
| 977.63 | 1.62 | 1.98 | 1.86 | 1.59 | 1.35 | 0.7 |
|  | 0.5 | 0.42 | 0.25 |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 |  |  |  |
| Q1swbeam | ffset |  |  |  |  |  |
| 2006202 |  |  |  |  |  |  |
| 110.1 |  |  |  |  |  |  |
| 127 |  |  |  |  |  |  |
| 1 | 0 | 20.9617 | 24.3417 | 10.5008 | 29.9494 | 15.518 |
|  | 15.7871 | 13.7063 | 12.2623 | 9.09681 | 8.07492 |  |
|  | 1.29966 | 3.32659 | 0.71346 | 1.24661 | 4.26735 |  |
|  | 1.11282 | 0.66165 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0.2503 | 17.4443 | 46.7889 | 15.0922 | 2.073 | 7.4772 |
|  | 2.9714 | 3.2638 | 8.2173 | 6.17255 | 2.70645 | 0 |
|  | 1.92946 | 1.46792 | 0.60148 | 0 | 1.04005 |  |
|  | 0.31768 | 0 | 0 | 0 | 0.50834 | 0 |
|  | 0 | 0.12415 | 0 | 0 |  |  |
| 1 | 0 | 11.9694 | 38.5327 | 44.2588 | 16.8101 |  |
|  | 17.9839 | 5.8217 | 8.2188 | 15.3684 | 6.56008 |  |
|  | 4.07772 | 3.63512 | 3.3578 | 2.4216 | 0.75185 | 0.2483 |
|  | 0.2483 | 0 | 5.56781 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 5.235 | 33.4672 | 22.5017 | 25.8667 |  |
|  | 12.4215 | 1.7615 | 3.6862 | 0.3434 | 7.53917 |  |
|  | 7.88458 | 1.66924 | 1.35582 | 0.26643 | 0.10484 | 0 |
|  | 0.10668 | 2.23934 | 3.64113 | 0 | 0 | 0 |
|  | 0 | 0 | 0.10455 | 1.76165 | 1.76165 |  |
| 1 | 0 | 23.8814 | 25.2679 | 44.1021 | 28.7077 |  |
|  | 18.1813 | 16.9984 | 1.0967 | 2.19 | 1.87814 |  |
|  | 3.87783 | 1.63973 | 1.19362 | 1.11282 | 0.70528 | 0 |
|  | 0.65966 | 0 | 0 | 0 | 1.11282 | 0 |
|  | 0 | 0 | 0 | 0 | 0 |  |
| 1 | 0 | 15.6093 | 31.6425 | 19.472 | 13.4385 | 6.9196 |
|  | 3.0124 | 4.3092 | 2.6587 | 0.38676 | 2.81983 |  |
|  | 1.21893 | 0.7916 | 0 | 0.30255 | 0.11824 | 0 |
|  | 0 | 0.17878 | 0 | 0 | 0.11824 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0 | 2.4032 | 35.864 | 28.1768 | 11.668 | 8.48 |
|  | 10.0964 | 7.2705 | 5.3933 | 0 | 0.20648 |  |
|  | 4.31641 | 0.67921 | 0 | 0.09932 | 0.65966 | 0 |
|  | 0.09932 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |
| 1 | 0 | 4.3218 | 17.2513 | 32.7718 | 26.2459 |  |
|  | 15.6648 | 12.9839 | 10.3966 | 9.301 | 7.65213 |  |
|  | 1.25131 | 7.2166 | 0 | 4.88124 | 1.36061 | 0 |

1

| 0 | 0 | 0 | 0 | 0 |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 7.49313 | 1.20296 | 0 | 0 | 0 | 0 |
| 1.2565 | 6.1443 | 38.1218 | 37.1957 | 10.5263 | 20.342 |
| 13.5544 | 26.5539 | 6.0348 | 2.87995 | 3.19119 |  |
| 1.29615 | 1.32124 | 2.31644 | 1.6562 | 2.0409 |  |
| 0.66062 | 0 | 0 | 2.0409 | 0.24598 | 0 |
| 0 | 0 | 0 | 0 | 0 |  |
| 0.9895 | 7.8397 | 16.9527 | 10.4441 | 12.6618 |  |
| 15.7945 | 22.0161 | 12.8365 | 11.1984 | 4.78045 |  |
| 4.96361 | 2.26539 | 7.19622 | 1.99517 | 0 |  |
| 4.59811 | 0.50271 | 1.08277 | 0.13153 | 0.38463 |  |
| 3.34157 | 0 | 0 | 0 | 0 | 0 |
| 0 |  |  |  |  |  |
| 0.12515 | 17.0147 | 37.47 | 16.1717 | 17.0353 |  |
| 12.0928 | 11.7792 | 15.8913 | 3.3377 | 3.79163 |  |
| 5.62407 | 3.36633 | 1.86454 | 0.12415 | 0 |  |
| 0.50124 | 0 | 0 | 0 | 1.67202 | 0 |
| 0.45985 | 0 | 0 | 0 | 0 | 0 |
| 1.38477 | 8.3704 | 21.7183 | 25.912 | 13.405 |  |
| 10.9208 | 8.4759 | 3.8099 | 2.7153 | 2.4744 | 0.5697 |
| 8.26016 | 0.86427 | 3.74239 | 0.16748 | 0.40391 | 0 |
| 0 | 0.33721 | 0.52773 | 0.19391 | 0 | 0 |
| 0 | 0 | 0 | 0 |  |  |
| 0 | 13.509 | 25.4432 | 38.4368 | 31.8088 |  |
| 12.9337 | 17.061 | 10.5022 | 8.6894 | 9.04688 |  |
| 2.42789 | 6.10779 | 5.71409 | 7.21492 | 1.67776 | 0 |
| 0 | 0 | 0 | 0.73992 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 |  |
| 0 | 23.4314 | 61.6211 | 29.4074 | 31.5671 |  |
| 17.6249 | 5.5207 | 3.9496 | 3.5844 | 2.57041 |  |
| 6.82601 | 0.1729 | 4.50089 | 1.29713 | 3.98425 | 0 |
| 0 | 0.39571 | 0 | 0.45154 | 0 | 0 |
| 0.83462 | 0 | 0 | 0 | 0 |  |
| 0.34121 | 21.5071 | 36.9707 | 39.7765 | 19.9786 |  |
| 14.7538 | 11.4951 | 5.0737 | 5.4513 | 4.28901 |  |
| 8.38046 | 0.94747 | 0.80528 | 0 | 0.18163 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0.2699 |
| 0 | 0 | 0 | 0 | 0 |  |
| 0.18772 | 6.7674 | 67.4654 | 39.6929 | 62.266 |  |
| 19.2075 | 16.213 | 25.5911 | 2.4385 | 6.22532 |  |
| 4.13235 | 0.28577 | 0.43425 | 3.3362 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0.14986 | 0 |
| 0.15319 | 0 | 0 | 0 | 0 |  |

FSP-UK
20032021
110.70 .75

127
$1 \quad 0.0005996525 \quad 0.1640287001 \quad 0.3331577428 \quad 0.3421042854 \quad 0.3077896855$
$\begin{array}{llllll}0.0276877607 & 0.0434349878 & 0.001199431 & 0.0606079973 & 0.0452031639\end{array}$
$\begin{array}{llllll}0.0762864358 & 0.0041900979 & 0.0044704097 & 0.0017318557 & 0.000259465\end{array}$
$\begin{array}{lllllll}0.0014588961 & 0 & 0 & 0.0001297325 & 0.0001297325 & 0\end{array}$
$\begin{array}{llllll}0.000259465 & 0.0001297325 & 0 & 0.0001297325 & 0 & 0\end{array}$
1
$\begin{array}{llllll}0.0001356596 & 0.1531649719 & 0.5482140297 & 0.3113148035 & 0.2625606101\end{array}$ $\begin{array}{llllll}0.1242854207 & 0.0562310285 & 0.0854244108 & 0.0348852879 & 0.0161248214\end{array}$ $\begin{array}{llllll}0.0169492456 & 0.0101294248 & 0.0064762205 & 0.0060502419 & 0.0044275753\end{array}$ $\begin{array}{llllll}0.0012155768 & 0.0011866052 & 0.0006639293 & 0.0034808758 & 0.0014061443\end{array}$ $\begin{array}{lllllll}0.0004170558 & 0.0013951526 & 0.0022790494 & 0.0007329086 & 0.0008245222 & 0\end{array}$ 0
1

| 0 | 0.1033295175 | 0.19641048 | 0.2419913717 | 0.1091266279 |
| :--- | :--- | :--- | :--- | :--- | :--- |

$\begin{array}{llllll}0.1568026119 & 0.145326301 & 0.036140277 & 0.0293963588 & 0.0143508007\end{array}$
$\begin{array}{llllll}0.0153718894 & 0.0071929565 & 0.0067527739 & 0.0018681385 & 0.009940521\end{array}$
$\begin{array}{llllll}0.00740716 & 0.0023788354 & 0.002716705 & 0.0021409311 & 0.0017422746\end{array}$ $\begin{array}{lllll}0.0005904058 & 0.0033955813 & 0.0006752624 & 0 & 0\end{array}$ 0.00023525980

| 0.0039650485 | 0.1530093905 | 0.338159914 | 0.1553922995 | 0.2150083739 |
| :--- | :--- | :--- | :--- | :--- | :--- |


| 0.0964601581 | 0.1165659253 | 0.1314649338 | 0.0263190435 | 0.0257133886 |
| :--- | :--- | :--- | :--- | :--- | :--- |


| 0.0180415374 | 0.0143487238 | 0.0090641849 | 0.0018065091 | 0.0041883778 |
| :--- | :--- | :--- | :--- | :--- | :--- |


| 0.0038052388 | 0.0035466212 | 0.0026256712 | 0.0019710043 | 0.0012706678 |
| :--- | :--- | :--- | :--- | :--- | :--- |

$0.00056455730 .00075812920 .0005729774 \quad 0.0012706678 \quad 3.38845 \mathrm{e}-05$ $3.38845 \mathrm{e}-050$
$\begin{array}{llllll}0.0010962171 & 0.1222741336 & 0.4498553047 & 0.2031795541 & 0.0752568916\end{array}$ $\begin{array}{llllll}0.0913421932 & 0.0585642238 & 0.0482046412 & 0.0990784093 & 0.0183456175\end{array}$ $\begin{array}{llllll}0.0234301884 & 0.0051412317 & 0.0112336251 & 0.0042618881 & 0.0040821708\end{array}$ $\begin{array}{llllll}0.0039301402 & 0 & 0.0003138761 & 0.0004997104 & 0.0009808918\end{array}$ $\begin{array}{llllll}0.0008212451 & 0.0011351212 & 0.001694755 & 0.0009721986 & 0 & 0\end{array}$ 0
$\begin{array}{llllll}2.26161 e-05 & 0.2195262683 & 0.3064700266 & 0.2655864135 & 0.2476267164\end{array}$ $\begin{array}{llllll}0.0431112703 & 0.0375114545 & 0.0146047595 & 0.0569714222 & 0.0329368056\end{array}$ $\begin{array}{llllll}0.0020584201 & 0.0103801271 & 0.0052184252 & 0.0003427702 & 0.0019322174\end{array}$ $\begin{array}{llllll}0.001240166 & 0.0002593175 & 0.0023037596 & 0.0017552985 & 0.0001146081\end{array}$ $\begin{array}{lllll}0.0019529181 & 0.0014618085 & 4.20676 e-05 & 0.0001036959 & 0.0016033953\end{array}$ 1.03851e-05 0

| 0 | 0.0871756844 | 0.2996241409 | 0.3111598691 | 0.1612888821 |
| :--- | :--- | :--- | :--- | :--- | :--- |


| 0.0607181418 | 0.0399573377 | 0.0280004615 | 0.0151930886 | 0.0179131136 |
| :--- | :--- | :--- | :--- | :--- | :--- | $\begin{array}{lllllll}0.047375509 & 0.0070657871 & 0.0029069767 & 0.0028085641 & 0.0034248136 & 0\end{array}$


| 0.0023009922 | 0 | 0 | 0 | 0.0014487729 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- |

0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: |

| 0 | 0.06241178 | 0.4007807363 | 0.0936432403 | 0.1661392336 |
| :--- | :--- | :--- | :--- | :--- |
| 0.084172186 | 0.0489667592 | 0.0045383432 | 0.0133242011 | 0.0072883556 |
| 0.010725752 | 0.0124001688 | 0.0133052176 | 0.0012450574 | 0.0037505906 |


| 7.88011e-05 | 0.0047453587 | 4.72807e-05 | 0.0012292972 | 1.57602e-05 |
| :---: | :---: | :---: | :---: | :---: |
| $3.15204 \mathrm{e}-05$ | $6.30409 \mathrm{e}-05$ | $3.15204 \mathrm{e}-05$ | 1.57602e-05 | 1.57602e-05 |
| 0.0034488814 | 0 |  |  |  |
| 0 | 0.0462429317 | 0.3661074051 | 0.3751123378 | 0.171327639 |
| 0.1173729434 | 0.0335259216 | 0.0444066234 | 0.0275603154 | 0.0030803347 |
| 0.006453779 | 0.0001320431 | 0.0090582307 | 0.0102501203 | 0.0064423865 |
| 0.006337584 | 0.0019806469 | 0 | 6.60216e-05 | $6.60216 \mathrm{e}-05$ |
| 0.0012992471 | 0.0001320431 | 6.60216e-05 | 0 | 6.60216e-05 |
| 0 |  |  |  |  |
| 0 | 0.0497881333 | 0.3584337435 | 0.4301705234 | 0.3611324055 |
| 0.1699642895 | 0.0915132656 | 0.0522974874 | 0.0372679269 | 0.0063585644 |
| 0.0155682804 | 0.0169229843 | 0.0150488511 | 0 | 0.0065329235 |
| 0.0065015519 | 0 | 0 | 0 | 0.0005752481 |
| 0 | 0 | 0 | 0 | 0 |
| 0 | 0.0992979309 | 0.3132769061 | 0.404824384 | 0.3187756661 |
| 0.2144234295 | 0.1202334108 | 0.0707920098 | 0.0346720214 | 0.0427286271 |
| 0.0019981129 | 0.0122453312 | 0.0035399024 | 0.0091491166 | 0.0040710437 |
| 0.0080827697 | 0.0039462729 | 0.0059770088 | 9.99056e-05 | 0.0015076265 |
| $9.99056 \mathrm{e}-05$ | 0 | 0 | $9.99056 \mathrm{e}-05$ | 0 |
| 0 |  |  |  |  |
| 0.0048896085 | 0.1093931233 | 0.2409909098 | 0.342222128 | 0.1843469436 |
| 0.1273706237 | 0.1082327865 | 0.0763853928 | 0.0576452317 | 0.0237495021 |
| 0.0250038238 | 0.00933254 | 0.0036693854 | 0.0111649666 | 0.0028229153 |
| 0.0031227752 | 0.0034365469 | 0.0047588703 | 0.0011504961 | 0 |
| 0 | 0 | 0 | 0.0014642678 | 0 |
| 0 | 0.106692296 | 0.462891223 | 0.1532642147 | 0.1442259017 |
| 0.1230641606 | 0.0781141924 | 0.102312786 | 0.0301074486 | 0.0478016471 |
| 0.014684173 | 0.0168037935 | 0.0050194129 | 0.0033997592 | 0.0021956008 |
| 0.0063508475 | 0.0048557028 | 0.003975375 | 0.0011243485 | 0.0027226531 |
| 0.0017564806 | 0 | 0 | 0 | 0 |
| 0 | 0.12886873 | 0.4887237516 | 0.3304347223 | 0.0960669305 |
| 0.0858467213 | 0.0726999591 | 0.0391128558 | 0.0589537549 | 0.0200598221 |
| 0.0122974024 | 0.0180049259 | 0.0119845876 | 0.0039815527 | 0.0027961056 |
| 0.0024247625 | 0.005117608 | 0.0059944095 | 0.0011243485 | 0 |
| 0 | 0 | 0 | 0 | 0 |
| 0 | 0.1586709782 | 0.2933003672 | 0.4153420564 | 0.2476762135 |
| 0.0907601391 | 0.0694807344 | 0.0922203076 | 0.0271016214 | 0.0610983845 |
| 0.0369478639 | 0.0380261073 | 0.0120135944 | 0.0073046846 | 0.0019174935 |
| 0 | 0.0011504961 | 0.0014381202 | 0.0026932068 | 0.0024055828 |
| 0 | 0 | 0.0011504961 | 0 | 0 |
| 0 | 0.2112724882 | 0.9077480144 | 0.1994578656 | 0.1528107317 |
| 0.0897784456 | 0.0471313352 | 0.0455976848 | 0.0374275951 | 0.0276157117 |
| 0.0403826099 | 0.0079840292 | 0.0096721581 | 0.0032406206 | 0.0017564806 |
| 0.0017564806 | 0.0028808291 | 0.0064020338 | 0 | 0.0045777045 |
| 0.00148414 | 0 | 0 | 0 | 0 |
| 0.001278329 | 0.3004545955 | 0.3368668308 | 0.4213025839 | 0.1902005191 |
| 0.1366906837 | 0.0642090392 | 0.0396030017 | 0.0304254687 | 0.0543559787 |
| 0.0184056185 | 0.042564176 | 0.0141117117 | 0.003992626 | 0.0064497077 |
| 0.0041487154 | 0.0024780937 | 0.0012243989 | 0.0001231713 | $2.46343 \mathrm{e}-05$ |
| 0.002276511 | 0.005323707 | $4.92685 \mathrm{e}-05$ | $2.46343 \mathrm{e}-05$ | $2.46343 \mathrm{e}-05$ |
| $2.46343 \mathrm{e}-05$ | 0 |  |  |  |
| 0 | 0.0457153066 | 0.3479819938 | 0.2646579215 | 0.3520245785 |
| 0.0947774833 | 0.0712985299 | 0.0333363482 | 0.0224343434 | 0.0129740757 |


| 0.0234239362 | 0 | 0.0232988167 | 0.0018886193 | 0.0018886193 |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 0.0064010685 | 0.0081597965 | 0.0029903139 | 0 | 0 |  |
| 0.0019875965 | 0.0018886193 | 0 | 0 | 0 | 0 | 0


| UK-CBT-ear |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19882002 |  |  |  |  |  |  |
| 1101 |  |  |  |  |  |  |
| 311 |  |  |  |  |  |  |
| 5.50 | 660.36 | 337.83 | 439.11 | 199.29 | 63.46 | 62.34 |
|  | 58.95 | 13.18 | 21.70 | 13.33 | 27.52 | 6.95 |
| 5.89 | 334.92 | 420.18 | 206.01 | 239.87 | 86.59 | 36.69 |
|  | 36.30 | 34.02 | 21.23 | 13.23 | 14.64 | 8.91 |
| 5.48 | 330.59 | 249.78 | 187.83 | 120.79 | 118.15 | 45.22 |
|  | 34.04 | 22.00 | 18.96 | 10.14 | 16.62 | 8.71 |
| 3.87 | 169.69 | 178.00 | 138.03 | 89.94 | 39.06 | 50.15 |
|  | 27.73 | 13.14 | 9.08 | 16.74 | 3.98 | 7.26 |
| 3.33 | 569.33 | 159.31 | 112.20 | 42.39 | 44.18 | 21.30 |
|  | 30.70 | 7.94 | 5.60 | 5.48 | 5.88 | 5.21 |
| 4.11 | 276.52 | 436.07 | 135.24 | 82.61 | 58.75 | 29.82 |
|  | 23.11 | 22.81 | 11.35 | 3.31 | 8.58 | 5.80 |
| 6.81 | 347.00 | 282.99 | 271.57 | 54.29 | 49.16 | 24.17 |
|  | 27.27 | 20.69 | 23.17 | 11.03 | 8.54 | 4.49 |
| 6.93 | 139.39 | 287.26 | 193.06 | 187.53 | 57.49 | 45.54 |
|  | 26.86 | 14.72 | 8.08 | 17.93 | 7.45 | 5.17 |
| 7.59 | 146.04 | 118.70 | 100.89 | 81.14 | 87.63 | 23.24 |
|  | 21.23 | 16.83 | 12.69 | 13.77 | 12.60 | 5.11 |
| 7.37 | 300.18 | 244.82 | 114.67 | 60.06 | 66.02 | 58.33 |
|  | 14.54 | 6.74 | 13.71 | 5.51 | 6.41 | 4.75 |
| 7.30 | 188.05 | 166.31 | 103.86 | 61.72 | 44.52 | 23.65 |
|  | 35.65 | 9.80 | 9.76 | 8.10 | 8.57 | 3.78 |
| 7.03 | 264.75 | 137.13 | 101.88 | 64.10 | 27.00 | 25.49 |
|  | 13.29 | 26.52 | 5.87 | 9.91 | 2.81 | 2.98 |
| 8.15 | 194.23 | 235.47 | 112.00 | 69.45 | 33.41 | 16.90 |
|  | 19.70 | 14.88 | 26.19 | 2.84 | 4.35 | 1.86 |
| 9.62 | 400.24 | 142.06 | 135.26 | 69.22 | 46.01 | 25.81 |
|  | 13.47 | 11.17 | 10.68 | 12.43 | 4.64 | 3.50 |
| 9.44 | 280.20 | 169.83 | 62.21 | 62.54 | 27.88 | 19.67 |
|  | 8.64 | 3.97 | 4.69 | 2.63 | 4.92 | 2.28 |



| 179.70 | 3.00 | 47.00 | 162.00 | 125.00 | 40.00 | 27.00 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 13.00 | 3.00 | 6.00 |  |  |  |
| 181.60 | 4.00 | 36.00 | 100.00 | 106.00 | 80.00 | 21.00 |
|  | 9.00 | 6.00 | 3.00 |  |  |  |

## Table 32.7. Sole in Division 7.e. Detailed XSA survivor diagnostics.

FLR XSA Diagnostics 2022-05-13 17:48:39
CPUE data from indices
Catch data for 53 years 1969 to 2021. Ages 2 to 12

|  | fleet first age | last age first year |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | UK-CBT-1ate | 3 | 11 | 2003 | 2021 | <NA> <NA> |
| 2 | UK-COT | 3 | 11 | 1988 | 2015 | <NA> <NA> |
| 3 | Q1SWBeam-nonoffset | 2 | 11 | 2006 | 2021 | <NA> <NA> |
| 4 | FSP-UK | 2 | 11 | 2004 | 2021 | <NA> <NA> |

Tapered time weighting applied Power = 3 over 15 years

Catchability analysis :
Catchability independent of size for all ages
Catchability independent of age for ages > 6
Terminal population estimation :
Survivor estimates shrunk towards the mean $F$
of the final 3 years or the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1$

Minimum standard error for population
estimates derived from each fleet $=0.4$
prior weighting not applied
Regression weights
year
age $20122013 \quad 20142015 \quad 2016 \quad 2017 \quad 2018 \quad 2019 \quad 20202021$
a11 0.4820 .610 .7250 .820 .8930 .9440 .9760 .9930 .999

Fishing mortalities year
$\begin{array}{llllllllllll}\text { age } & 2012 & 2013 & 2014 & 2015 & 2016 & 2017 & 2018 & 2019 & 2020 & 2021\end{array}$
$2 \quad 0.009 \quad 0.039 \quad 0.062 \quad 0.046 \quad 0.0170 .027 \quad 0.009 \quad 0.0190 .0200 .045$
$3 \quad 0.081 \quad 0.113 \quad 0.125 \quad 0.125 \quad 0.131 \quad 0.276 \quad 0.115 \quad 0.1190 .1090 .117$
$4 \quad 0.175 \quad 0.2220 .212 \quad 0.201 \quad 0.235 \quad 0.1310 .1290 .2210 .2120 .173$
$\begin{array}{llllllllllllllllllll}5 & 0.247 & 0.273 & 0.278 & 0.209 & 0.189 & 0.213 & 0.135 & 0.269 & 0.294 & 0.299\end{array}$
$\begin{array}{llllllllllllllllll}6 & 0.291 & 0.244 & 0.273 & 0.244 & 0.211 & 0.261 & 0.342 & 0.253 & 0.273 & 0.438\end{array}$
$\begin{array}{lllllllllllll}7 & 0.305 & 0.255 & 0.203 & 0.191 & 0.333 & 0.145 & 0.207 & 0.306 & 0.239 & 0.297\end{array}$
$\begin{array}{llllllllllll}8 & 0.256 & 0.289 & 0.184 & 0.186 & 0.241 & 0.187 & 0.388 & 0.318 & 0.319 & 0.227\end{array}$
$\begin{array}{lllllllllllllllllll}9 & 0.263 & 0.261 & 0.221 & 0.170 & 0.140 & 0.068 & 0.271 & 0.309 & 0.229 & 0.462\end{array}$
100.3590 .2730 .2310 .1680 .1400 .0940 .1670 .2790 .2670 .322
$\begin{array}{lllllllllllll}11 & 0.392 & 0.284 & 0.216 & 0.153 & 0.163 & 0.559 & 1.108 & 0.237 & 0.184 & 0.257\end{array}$
$\begin{array}{llllllllllllllllll}12 & 0.392 & 0.284 & 0.216 & 0.153 & 0.163 & 0.559 & 1.108 & 0.237 & 0.184 & 0.257\end{array}$

XSA population number (Thousand) age

|  |  |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| 2012 | 3540 | 3054 | 3449 | 1924 | 1478 | 923 | 638 | 306 | 153 | 125 | 401 |
| 2013 | 3282 | 3174 | 2547 | 2621 | 1360 | 1000 | 615 | 447 | 213 | 96 | 467 |
| 2014 | 3464 | 2856 | 2564 | 1845 | 1805 | 964 | 701 | 417 | 312 | 147 | 524 |
| 2015 | 4160 | 2946 | 2279 | 1876 | 1265 | 1243 | 712 | 528 | 303 | 224 | 610 |
| 2016 | 5817 | 3596 | 2353 | 1687 | 1378 | 896 | 929 | 535 | 403 | 231 | 750 |
| 2017 | 4908 | 5176 | 2854 | 1683 | 1264 | 1010 | 581 | 661 | 421 | 317 | 190 |
| 2018 | 9132 | 4324 | 3554 | 2265 | 1231 | 880 | 790 | 436 | 559 | 347 | 240 |
| 2019 | 6378 | 8188 | 3488 | 2825 | 1790 | 791 | 648 | 485 | 301 | 428 | 781 |
| 2020 | 9050 | 5662 | 6577 | 2529 | 1954 | 1258 | 527 | 426 | 322 | 206 | 992 |
| 2021 | 2745 | 8028 | 4592 | 4815 | 1706 | 1346 | 896 | 347 | 307 | 223 | 824 |

Estimated population abundance at 1st Jan 2022

| year | $\begin{array}{l}\text { age } \\ 2\end{array}$ |  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

2022023746464349732329969056461981201156

Fleet: UK-CBT-1ate
Log catchability residuals.

| age | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | -0.274 | -0.487 | 0.571 | 0.763 | 0.884 | 0.541 | 0.244 | -0.597 | 0.128 | -0.293 |
| 4 | -0.577 | -1.221 | 0.218 | 0.284 | 0.280 | 0.236 | 0.039 | -0.160 | -0.417 | -0.096 |
| 5 | -0.812 | -1.417 | -0.087 | 0.311 | 0.102 | -0.081 | 0.106 | 0.043 | -0.060 | -0.398 |
| 6 | -1.696 | -0.973 | 0.065 | -0.084 | -0.028 | 0.075 | 0.169 | -0.069 | -0.065 | -0.135 |
| 7 | -1.319 | -2.014 | 0.117 | 0.155 | -0.020 | 0.040 | -0.276 | -0.039 | -0.089 | -0.108 |
| 8 | -1.393 | -1.175 | -0.689 | 0.357 | 0.032 | 0.266 | -0.164 | -0.036 | -0.192 | -0.086 |
| 9 | -0.722 | -1.379 | 0.130 | 0.031 | 0.072 | -0.055 | -0.227 | -0.026 | -0.306 | -0.123 |
| 10 | -1.249 | -1.062 | 0.197 | 0.199 | -0.411 | 0.119 | -0.400 | -0.427 | 0.136 | 0.079 |
| 11 | -1.345 | -1.219 | 0.370 | 0.155 | 0.108 | -0.027 | -0.340 | -0.083 | 0.088 | 0.113 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |  |
| 3 | -0.665 | -0.244 | -0.130 | 0.237 | 0.399 | -0.175 | 0.242 | 0.124 | 0.080 |  |
| 4 | 0.102 | -0.096 | -0.164 | -0.071 | 0.084 | 0.175 | 0.146 | 0.034 | -0.012 |  |
| 5 | 0.148 | -0.032 | 0.019 | -0.198 | -0.162 | -0.053 | 0.112 | 0.259 | 0.122 |  |
| 6 | -0.029 | 0.104 | -0.005 | -0.021 | -0.129 | -0.204 | 0.013 | 0.042 | 0.313 |  |
| 7 | -0.083 | -0.001 | -0.160 | 0.158 | -0.126 | 0.083 | 0.128 | 0.062 | 0.187 |  |
| 8 | 0.164 | -0.085 | -0.184 | -0.073 | 0.059 | -0.157 | -0.084 | 0.372 | -0.066 |  |
| 9 | -0.022 | 0.021 | -0.250 | -0.303 | -0.240 | -0.016 | 0.005 | -0.108 | 0.284 |  |
| 10 | 0.137 | -0.041 | -0.063 | -0.101 | -0.323 | -0.357 | 0.028 | -0.248 | 0.185 |  |
| 11 | 0.108 | 0.054 | -0.203 | -0.288 | -0.234 | 0.159 | -0.235 | -0.585 | 0.050 |  |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean_Logq | -5.0580 | -4.4680 | -4.3435 | -4.2893 | -4.2893 | -4.2893 | -4.2893 |

Fleet: UK-COT
Log catchability residuals.

|  | year $1988$ | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 1.117 | 0.757 | 1.029 | 0.755 | 0.721 | 0.679 | 0.912 | 0.357 | 0.798 | 1.129 |
| 4 | 0.647 | 0.463 | 0.461 | 0.007 | -0.036 | 0.202 | 0.456 | 0.537 | 0.525 | 0.336 |
| 5 | 0.537 | 0.417 | 0.303 | -0.259 | -0.418 | 0.272 | 0.201 | 0.192 | 0.310 | 0.352 |
| 6 | 0.358 | 0.316 | 0.476 | -0.554 | -1.335 | -0.099 | -0.111 | -0.234 | 0.027 | -0.007 |
| 7 | -0.345 | 0.173 | 0.084 | -0.824 | -0.977 | -0.192 | 0.096 | 0.020 | -0.202 | 0.230 |
| 8 | -0.154 | -0.096 | -0.105 | -0.780 | -1.196 | -0.463 | -0.495 | -0.051 | -0.012 | -0.366 |
| 9 | -0.536 | 0.147 | 0.112 | -1.141 | -1.299 | 0.012 | 0.022 | -0.496 | -0.080 | -0.209 |
| 10 | -1.262 | -0.142 | -0.072 | -0.854 | -1.728 | -0.571 | 0.606 | -0.596 | -0.235 | -0.915 |
| 11 | -0.372 | 0.422 | -0.437 | -1.111 | -2.082 | -0.130 | 0.087 | -0.333 | -0.185 | -0.508 |
| age | year 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| 3 | 0.796 | 0.894 | 0.465 | 0.838 | 0.517 | 0.251 | 0.170 | 1.150 | 0.856 | 0.937 |
| 4 | 0.344 | 0.243 | 0.322 | -0.035 | -0.359 | -0.091 | -0.573 | 0.729 | 0.306 | 0.326 |
| 5 | 0.026 | 0.285 | 0.090 | 0.132 | -0.432 | -0.278 | -0.786 | 0.313 | 0.288 | 0.181 |
| 6 | 0.169 | -0.003 | -0.110 | 0.015 | 0.019 | -0.968 | -0.385 | 0.388 | -0.252 | 0.001 |
| 7 | -0.090 | 0.192 | -0.557 | 0.042 | -0.156 | -0.636 | -1.491 | 0.418 | -0.068 | -0.011 |
| 8 | -0.383 | 0.428 | -0.377 | -0.350 | -0.276 | -0.683 | -0.714 | -0.351 | 0.218 | 0.075 |
| 9 | -0.485 | -0.199 | 0.288 | 0.189 | -1.054 | -0.186 | -0.867 | 0.490 | -0.123 | 0.081 |
| 10 | -0.435 | -0.055 | -0.091 | 0.381 | -0.471 | -0.702 | -0.549 | 0.487 | -0.027 | -0.335 |
| 11 | -0.040 | 0.281 | -0.161 | 0.118 | 0.324 | -0.416 | -0.656 | 0.865 | -0.097 | -0.033 |
| age | $\text { year } 2008$ | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |  |  |
| 3 | 0.609 | 0.812 | -0.394 | 0.341 | 0.171 | -0.333 | -0.067 | 0.015 |  |  |
| 4 | 0.162 | 0.246 | -0.021 | -0.406 | 0.250 | 0.141 | -0.016 | -0.106 |  |  |
| 5 | -0.195 | 0.146 | 0.091 | -0.131 | -0.105 | 0.215 | 0.030 | -0.106 |  |  |
| 6 | -0.091 | 0.073 | -0.066 | -0.243 | 0.069 | -0.027 | 0.091 | 0.015 |  |  |
| 7 | -0.064 | -0.357 | -0.028 | -0.277 | -0.022 | -0.139 | 0.003 | -0.157 |  |  |
| 8 | 0.145 | -0.054 | -0.001 | -0.587 | 0.057 | 0.207 | -0.021 | -0.259 |  |  |
| 9 | -0.109 | -0.127 | 0.028 | -0.516 | 0.032 | -0.002 | 0.156 | -0.304 |  |  |
| 10 | 0.083 | -0.208 | -0.265 | -0.146 | 0.257 | 0.283 | 0.173 | 0.078 |  |  |
| 11 | 0.069 | -0.072 | 0.062 | -0.234 | 0.410 | 0.312 | 0.227 | -0.147 |  |  |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean_Logq | -14.2957 | -13.6814 | -13.5450 | -13.4117 | -13.4117 | -13.4117 | -13.4117 |
| S.E_Logq | 0.4270 | 0.3152 | 0.2995 | 0.3768 | 0.4016 | 0.3552 | 0.4393 |
| Mean_Logq | -13.4117 | -13.4117 |  |  |  |  |  |
| S.E_Logq | 0.5235 | 0.5434 |  |  |  |  |  |

Fleet: Q1SWBeam-nonoffset
Log catchability residuals.

| age | $2006$ | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.790 | 0.754 | 0.313 | -0.413 | 0.849 | 0.764 | -1.130 | -0.463 | -0.161 | -0.103 |
| 3 | 0.015 | 0.575 | 0.472 | 0.265 | 0.054 | -0.002 | 0.466 | -0.299 | 0.602 | -0.240 |
| 4 | -0.205 | -0.222 | 0.740 | 0.123 | 0.649 | -0.137 | -0.046 | 0.416 | 0.535 | -0.620 |
| 5 | 0.622 | -1.435 | 0.304 | 0.526 | 0.571 | -0.352 | -0.423 | 0.084 | -0.478 | -0.322 |
| 6 | 0.640 | -0.275 | 1.153 | 0.384 | 0.457 | -0.580 | -0.519 | 0.170 | 0.153 | 0.251 |
| 7 | 0.672 | -0.536 | 0.014 | -0.755 | 1.071 | -1.038 | 0.129 | 0.292 | 0.362 | 0.591 |
| 8 | 0.743 | -0.485 | 1.009 | 0.016 | -0.928 | 0.046 | 0.161 | 0.561 | 1.350 | 0.608 |
| 9 | 1.365 | 0.661 | 1.440 | -1.675 | -0.154 | 0.261 | 0.599 | 0.764 | 0.394 | 0.767 |
| 10 | 1.787 | 1.098 | 0.834 | 1.050 | 0.347 | -1.477 | 0.000 | 1.314 | -0.052 | 0.473 |
| 11 | 2.061 | 1.105 | 1.106 | 1.380 | 0.699 | 1.056 | -1.745 | 0.296 | 0.802 | 0.809 |
| age | year 2016 | 2017 | 2018 | 2019 | 2020 | 2021 |  |  |  |  |
| 2 | 0.331 | -0.206 | -0.352 | 0.560 | 0.124 | 0.165 |  |  |  |  |
| 3 | 0.355 | -0.529 | -0.219 | 0.028 | -0.116 | 0.137 |  |  |  |  |
| 4 | -0.208 | 0.052 | 0.227 | -0.006 | -0.340 | 0.010 |  |  |  |  |
| 5 | 0.077 | -0.156 | 0.398 | 0.192 | -0.150 | 0.344 |  |  |  |  |
| 6 | -0.108 | -0.114 | 0.095 | 0.015 | -0.247 | 0.181 |  |  |  |  |
| 7 | 0.317 | -0.163 | 0.684 | -0.320 | -0.062 | 0.224 |  |  |  |  |
| 8 | 0.565 | -0.403 | 0.338 | -0.453 | 0.004 | 1.075 |  |  |  |  |
| 9 | -0.462 | -0.892 | 0.723 | -0.262 | 0.272 | -0.285 |  |  |  |  |
| 10 | -0.051 | -0.529 | 0.497 | -0.123 | 0.320 | 0.750 |  |  |  |  |
| 11 | 0.902 | -1.634 | -0.178 | 0.495 | 1.421 | 0.648 |  |  |  |  |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

|  | 2 |  | 3 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean_Logq | -6.1455 | -4.8787 | -4.7133 | -4.6222 | -4.5738 | -4.5738 | -4.5738 | -4.5738 |
| S.E_Logq | 0.5609 | 0.3354 | 0.3737 | 0.5259 | 0.4358 | 0.5615 | 0.6352 | 0.8140 |
|  | 10 | 11 |  |  |  |  |  |  |
| Mean_Logq | -4.5738 | -4.5738 |  |  |  |  |  |  |
| S.E_Logq | 0.7844 | 1.0169 |  |  |  |  |  |  |

Fleet: FSP-UK
Log catchability residuals.

| age | $e_{2 r}{ }_{2004}$ | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.961 | 0.173 | 0.459 | 0.346 | 0.876 | 0.040 | 0.084 | -0.221 | -0.554 | -0.383 | 0.270 |
| 3 | 0.537 | 0.132 | 0.280 | 0.469 | 0.171 | 0.031 | -0.349 | 0.086 | 0.312 | 0.276 | 0.255 |
| 4 | 0.578 | 0.063 | 0.239 | 0.149 | 0.252 | 0.393 | -0.002 | -0.912 | 0.177 | 0.652 | 0.577 |
| 5 | 0.485 | -0.003 | 0.467 | 0.001 | 0.815 | 0.120 | 0.168 | -0.077 | 0.033 | 0.488 | 0.718 |
| 6 | 0.011 | 0.486 | 0.4390 | 0.247 | -0.015 | -0.106 | 0.284 | -0.192 | 0.039 | 0.459 | 0.429 |
| 7 | 0.040 | 0.622 | 0.622 | 0.451 | -0.141 | 0.248 | -0.278 | -0.363 | -0.732 | 0.155 | 0.427 |
| 8 | 1.039 | -0.018 | 0.959 | 0.136 | -0.412 | -0.057 | -0.096 | -2.010 | -0.118 | 0.106 | 0.202 |
| 9 | 0.618 | 0.446 | 0.076 | 1.088 | 0.644 | 0.014 | 0.316 | -0.237 | 0.146 | 0.066 | 0.034 |
| 10 | 0.431 | 0.152 | 0.842 | 0.133 | 0.370 | -0.195 | 0.449 | -0.643 | -1.281 | -0.951 | 0.542 |
| 11 | 0.633 | 0.855 | 0.834 | 1.228 | -1.616 | 1.055 | 0.381 | 0.349 | -0.316 | 0.000 | -1.776 |
| age | $\begin{aligned} & \text { year } \\ & 2015 \end{aligned}$ | 2016 | - 2017 |  | 18201 | 202 | 020 | 021 |  |  |  |
| 2 | 0.172 | -0.210 | 0.157 | $7-0.2$ | $69 \quad 0.3$ | 3830.3 | 386-0. | 285 |  |  |  |
| 3 | -0.039 | 0.419 | 0.214 | -0.2 | 330.2 | 261-0.3 | 368-0. | . 680 |  |  |  |
| 4 | 0.519 | -0.292 | 0.208 |  | $16-0.4$ | 432-0.3 | 325-0. | 459 |  |  |  |
| 5 | 0.104 | -0.050 | -0.436 | 6 0.15 | $57-0.4$ | 450-0.1 | $102-0$. | 127 |  |  |  |
| 6 | 0.243 | 0.099 | -0.138 | 80.0 | -02-0.4 | 448-0.1 | 101-0. | , 211 |  |  |  |
| 7 | 0.059 | 0.163 | -0.164 | -0.020 | $28-0.23$ | 237-0.4 | 440-0. | . 361 |  |  |  |
| 8 | 0.264 | 0.330 | -0.201 | 10.4 | $95-0.0$ | 0610.00 | 004-0.7 | . 765 |  |  |  |
| 9 | 0.270 | -0.415 | -0.006 | -0.2 |  | . 175 -0.1 | 112-0.0 | . 041 |  |  |  |
| 10 | -0.061 | 0.331 | -0.614 | 40.2 | 70 0.1 | 1750.7 | 777 -0. | . 568 |  |  |  |
| 11 | 0.281 | -0.278 | -0.483 | 30.9 | 260.1 | 1730.0 | 0790. | 295 |  |  |  |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean_Logq | -10.6122 | -9.2096 | -9.1044 | -9.1078 | -9.1971 | -9.1971 | -9.1971 | -9.1971 |
| S.E_Logq | 0.4107 | 0.3227 | 0.4243 | 0.3499 | 0.2688 | 0.3795 | 0.6578 | 0.3652 |
|  | 10 | 11 |  |  |  |  |  |  |
| Mean_Logq | -9.1971 | -9.1971 |  |  |  |  |  |  |
| S.E_Logq | 0.5981 | 0.8250 |  |  |  |  |  |  |

Terminal year survivor and $F$ summaries:
, Age 2 Year class =2019

| source | scaledWts | survivors | yrc1s |
| :--- | ---: | ---: | ---: |
|  | 0.362 | 2801 | 2019 |
| Q1SWBeam-nonoffset | 0.546 | 1785 | 2019 |

fshk $0.091 \quad 6778 \quad 2019$
, Age 3 Year class $=2018$

| source |  |  |  |
| :--- | ---: | ---: | ---: |
|  | scaledWts | survivors | yrcls |
| UK-CBT-7ate | 0.314 | 7000 | 2018 |
| Q1SWBeam-nonoffset | 0.314 | 7417 | 2018 |
| FSP-UK | 0.314 | 3276 | 2018 |
| fshk | 0.057 | 6589 | 2018 |

, Age 4 Year class $=2017$
source

|  | scaledwts | survivors | yrcls |
| :--- | ---: | ---: | ---: |
|  | 0.347 | 3455 | 2017 |
| UK-CBT-7ate | 0.347 | 3531 | 2017 |
| Q1SWBeam-nonoffset | 0.239 | 2209 | 2017 |
| FSP-UKK | 0.066 | 3188 | 2017 |

,Age 5 Year class $=2016$
source

|  | scaledwts | survivors | yrcls |
| :--- | ---: | ---: | ---: |
| UK-CBT-7ate | 0.311 | 3651 | 2016 |
| Q1SWBeam-nonoffset | 0.311 | 4558 | 2016 |
| FSP-UK | 0.311 | 2847 | 2016 |
| fshk | 0.067 | 4285 | 2016 |

,Age 6 Year class $=2015$
source

| source | scaledwts | survivors | yrcls |
| :--- | ---: | ---: | ---: |
| UK-CBT-1ate | 0.308 | 1361 | 2015 |
| Q1SWBeam-nonoffset | 0.308 | 1194 | 2015 |
| FSP-UK | 0.308 | 807 | 2015 |
| fshk | 0.076 | 1628 | 2015 |

,Age 7 Year class =2014
source

|  | scaledwts | survivors | yrcls |
| :--- | ---: | ---: | ---: |
|  | 0.351 | 1092 | 2014 |
| UK-CBT-7ate | 0.222 | 1132 | 2014 |
| Q1SWBeam-nonoffset | 0.351 | 631 | 2014 |
| FSP-UK | 0.076 | 1095 | 2014 |

,Age 8 Year class =2013

| source |  | scaledwts | survivors | yrcls |
| :--- | ---: | ---: | ---: | ---: |
|  | 0.507 | 605 | 2013 |  |
| UK-CBT-late | 0.013 |  |  |  |
| Q1SWBeam-nonoffset | 0.150 | 1894 | 2013 |  |
| FSP-UK | 0.241 | 301 | 2013 |  |
| fshk | 0.102 | 402 | 2013 |  |

,Age 9 Year class =2012

| source |  | scaledwts | survivors |
| :--- | ---: | ---: | ---: |
| Urcls |  |  |  |
| UK-CBT-late | 0.382 | 263 | 2012 |
| Q1SWBeam-nonoffset | 0.139 | 149 | 2012 |
| FSP-UK | 0.382 | 190 | 2012 |
| fshk | 0.097 | 374 | 2012 |

,Age 10 Year class =2011

| source |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
|  | scaledwts | survivors | yrcls |
| UK-CBT-1ate | 0.523 | 242 | 2011 |
| Q1SWBeam-nonoffset | 0.171 | 426 | 2011 |
| FSP-UK | 0.191 | 114 | 2011 |
| fshk | 0.116 | 284 | 2011 |

,Age 11 Year class =2010

| source |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
|  | scaledwts | survivors | yrcls |
| UK-CBT-late | 0.622 | 164 | 2010 |
| Q1SWBeam-nonoffset | 0.079 | 298 | 2010 |
| FSP-UK | 0.170 | 210 | 2010 |
| fshk | 0.129 | 109 | 2010 |

Table 32.8. Sole in Division 7.e. Estimated stock numbers-at-age (thousands).

| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 1874 | 2380 | 625 | 966 | 1513 | 159 | 507 | 572 | 262 | 90 | 636 | 9585 |
| 1970 | 1343 | 1611 | 1848 | 490 | 732 | 1170 | 124 | 412 | 494 | 218 | 1123 | 9564 |
| 1971 | 3826 | 1164 | 1237 | 1365 | 358 | 584 | 952 | 100 | 340 | 397 | 821 | 11144 |
| 1972 | 2568 | 3414 | 863 | 885 | 1047 | 262 | 452 | 713 | 81 | 274 | 542 | 11102 |
| 1973 | 2264 | 2185 | 2698 | 621 | 691 | 840 | 224 | 386 | 518 | 37 | 1222 | 11687 |
| 1974 | 3107 | 1981 | 1600 | 2029 | 478 | 532 | 646 | 187 | 300 | 440 | 850 | 12149 |
| 1975 | 2967 | 2769 | 1461 | 1238 | 1667 | 365 | 406 | 544 | 138 | 248 | 1756 | 13559 |
| 1976 | 2791 | 2606 | 1966 | 1160 | 931 | 1399 | 304 | 317 | 468 | 105 | 1598 | 13645 |
| 1977 | 6556 | 2367 | 1960 | 1330 | 896 | 714 | 1178 | 230 | 231 | 375 | 1866 | 17703 |
| 1978 | 4657 | 5527 | 1839 | 1408 | 1007 | 714 | 580 | 995 | 199 | 186 | 1385 | 18497 |
| 1979 | 4389 | 3976 | 3933 | 1334 | 1070 | 732 | 547 | 456 | 827 | 144 | 1493 | 18901 |
| 1980 | 4702 | 3755 | 2834 | 2787 | 970 | 751 | 497 | 397 | 327 | 650 | 1702 | 19372 |
| 1981 | 8130 | 4088 | 2866 | 2091 | 1923 | 758 | 506 | 316 | 298 | 243 | 934 | 22154 |
| 1982 | 4679 | 7124 | 2932 | 1974 | 1448 | 1370 | 516 | 337 | 214 | 214 | 1035 | 21844 |
| 1983 | 3866 | 4113 | 5065 | 1782 | 1260 | 976 | 1011 | 337 | 198 | 117 | 828 | 19554 |
| 1984 | 5968 | 3412 | 3005 | 3087 | 1058 | 806 | 629 | 635 | 192 | 110 | 982 | 19884 |
| 1985 | 6982 | 5083 | 2456 | 1934 | 2073 | 648 | 535 | 446 | 430 | 123 | 532 | 21242 |
| 1986 | 3765 | 6044 | 2982 | 1504 | 1303 | 1319 | 417 | 392 | 306 | 309 | 529 | 18870 |
| 1987 | 5847 | 3173 | 3930 | 1774 | 961 | 874 | 874 | 297 | 283 | 191 | 754 | 18959 |
| 1988 | 3878 | 4827 | 2102 | 2519 | 1199 | 675 | 578 | 573 | 224 | 208 | 713 | 17497 |
| 1989 | 3735 | 3088 | 3000 | 1335 | 1587 | 729 | 465 | 369 | 415 | 166 | 743 | 15632 |
| 1990 | 2817 | 3009 | 1965 | 1542 | 736 | 952 | 445 | 316 | 232 | 268 | 739 | 13022 |
| 1991 | 7161 | 2225 | 1864 | 1225 | 870 | 434 | 610 | 267 | 189 | 139 | 656 | 15639 |
| 1992 | 3902 | 6051 | 1618 | 1228 | 833 | 578 | 304 | 446 | 177 | 136 | 528 | 15801 |
| 1993 | 3350 | 3230 | 4111 | 1068 | 829 | 645 | 416 | 217 | 334 | 136 | 344 | 14679 |
| 1994 | 2378 | 2832 | 2253 | 2667 | 633 | 541 | 440 | 302 | 140 | 249 | 487 | 12921 |
| 1995 | 3452 | 2059 | 1938 | 1508 | 1882 | 466 | 389 | 351 | 218 | 84 | 646 | 12994 |
| 1996 | 3940 | 3033 | 1570 | 1155 | 958 | 1312 | 297 | 256 | 259 | 166 | 649 | 13597 |
| 1997 | 3331 | 3218 | 2322 | 1074 | 762 | 644 | 943 | 204 | 174 | 188 | 412 | 13271 |


| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 4398 | 2808 | 2121 | 1412 | 662 | 518 | 398 | 689 | 143 | 138 | 648 | 13936 |
| 1999 | 3585 | 3727 | 1965 | 1409 | 958 | 400 | 325 | 284 | 502 | 97 | 478 | 13730 |
| 2000 | 6522 | 2977 | 2503 | 1302 | 896 | 624 | 253 | 197 | 206 | 352 | 366 | 16198 |
| 2001 | 5423 | 5610 | 2124 | 1550 | 829 | 593 | 462 | 179 | 114 | 138 | 515 | 17538 |
| 2002 | 3815 | 4769 | 3743 | 1417 | 930 | 496 | 367 | 323 | 110 | 62 | 286 | 16319 |
| 2003 | 5381 | 3136 | 3125 | 2586 | 914 | 535 | 326 | 232 | 203 | 68 | 303 | 16809 |
| 2004 | 2870 | 4300 | 2043 | 1921 | 1726 | 703 | 413 | 248 | 155 | 124 | 298 | 14801 |
| 2005 | 3994 | 2218 | 2863 | 1423 | 1315 | 1062 | 480 | 264 | 166 | 89 | 351 | 14225 |
| 2006 | 4612 | 3369 | 1562 | 1797 | 860 | 842 | 682 | 327 | 162 | 108 | 301 | 14622 |
| 2007 | 3934 | 3698 | 2300 | 964 | 1050 | 540 | 549 | 441 | 215 | 94 | 348 | 14133 |
| 2008 | 4209 | 3369 | 2536 | 1363 | 594 | 605 | 318 | 373 | 295 | 141 | 319 | 14123 |
| 2009 | 3788 | 3541 | 2333 | 1650 | 876 | 383 | 373 | 188 | 270 | 203 | 374 | 13978 |
| 2010 | 4864 | 3269 | 2690 | 1744 | 1176 | 600 | 283 | 267 | 135 | 197 | 457 | 15683 |
| 2011 | 3471 | 4337 | 2627 | 2059 | 1265 | 870 | 422 | 210 | 174 | 103 | 338 | 15876 |
| 2012 | 3540 | 3054 | 3449 | 1924 | 1478 | 923 | 638 | 306 | 153 | 125 | 401 | 15991 |
| 2013 | 3282 | 3174 | 2547 | 2621 | 1360 | 1000 | 615 | 447 | 213 | 96 | 467 | 15823 |
| 2014 | 3464 | 2856 | 2564 | 1845 | 1805 | 964 | 701 | 417 | 312 | 147 | 524 | 15599 |
| 2015 | 4160 | 2946 | 2279 | 1876 | 1265 | 1243 | 712 | 528 | 303 | 224 | 610 | 16148 |
| 2016 | 5817 | 3596 | 2353 | 1687 | 1378 | 896 | 929 | 535 | 403 | 231 | 750 | 18577 |
| 2017 | 4908 | 5176 | 2854 | 1683 | 1264 | 1010 | 581 | 661 | 421 | 317 | 190 | 19065 |
| 2018 | 9132 | 4324 | 3554 | 2265 | 1231 | 880 | 790 | 436 | 559 | 347 | 240 | 23758 |
| 2019 | 6378 | 8188 | 3488 | 2825 | 1790 | 791 | 648 | 485 | 301 | 428 | 781 | 26103 |
| 2020 | 9050 | 5662 | 6577 | 2529 | 1954 | 1258 | 527 | 426 | 322 | 206 | 992 | 29504 |
| 2021 | 2745 | 8028 | 4592 | 4815 | 1706 | 1346 | 896 | 347 | 307 | 223 | 824 | 25829 |

Table 32.9. Sole in Division 7.e. Estimated fishing mortality-at-age.

| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | $F_{\text {bar }}$ <br> (3-9) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 0.051 | 0.153 | 0.144 | 0.176 | 0.157 | 0.151 | 0.108 | 0.048 | 0.084 | 0.110 | 0.110 | 0.134 |
| 1970 | 0.043 | 0.164 | 0.203 | 0.213 | 0.126 | 0.106 | 0.115 | 0.093 | 0.118 | 0.112 | 0.112 | 0.146 |
| 1971 | 0.014 | 0.200 | 0.234 | 0.165 | 0.212 | 0.155 | 0.188 | 0.109 | 0.113 | 0.156 | 0.156 | 0.181 |
| 1972 | 0.062 | 0.136 | 0.228 | 0.147 | 0.120 | 0.059 | 0.059 | 0.219 | 0.690 | 0.230 | 0.230 | 0.138 |
| 1973 | 0.034 | 0.212 | 0.185 | 0.163 | 0.163 | 0.162 | 0.081 | 0.152 | 0.063 | 0.124 | 0.124 | 0.160 |
| 1974 | 0.015 | 0.205 | 0.156 | 0.097 | 0.170 | 0.171 | 0.072 | 0.199 | 0.089 | 0.140 | 0.140 | 0.153 |
| 1975 | 0.029 | 0.243 | 0.130 | 0.185 | 0.075 | 0.083 | 0.147 | 0.051 | 0.181 | 0.108 | 0.108 | 0.131 |
| 1976 | 0.065 | 0.185 | 0.291 | 0.158 | 0.166 | 0.072 | 0.176 | 0.216 | 0.122 | 0.151 | 0.151 | 0.180 |
| 1977 | 0.071 | 0.152 | 0.230 | 0.178 | 0.128 | 0.108 | 0.069 | 0.048 | 0.114 | 0.093 | 0.093 | 0.130 |
| 1978 | 0.058 | 0.240 | 0.221 | 0.174 | 0.220 | 0.165 | 0.140 | 0.085 | 0.226 | 0.167 | 0.167 | 0.178 |
| 1979 | 0.056 | 0.239 | 0.244 | 0.219 | 0.254 | 0.287 | 0.221 | 0.232 | 0.142 | 0.228 | 0.228 | 0.242 |
| 1980 | 0.040 | 0.170 | 0.204 | 0.271 | 0.147 | 0.295 | 0.352 | 0.188 | 0.198 | 0.236 | 0.236 | 0.232 |
| 1981 | 0.032 | 0.232 | 0.273 | 0.268 | 0.239 | 0.285 | 0.305 | 0.289 | 0.229 | 0.270 | 0.270 | 0.270 |
| 1982 | 0.029 | 0.241 | 0.398 | 0.349 | 0.295 | 0.203 | 0.325 | 0.434 | 0.503 | 0.353 | 0.353 | 0.321 |
| 1983 | 0.025 | 0.214 | 0.395 | 0.421 | 0.347 | 0.340 | 0.365 | 0.462 | 0.489 | 0.402 | 0.402 | 0.363 |
| 1984 | 0.061 | 0.229 | 0.341 | 0.298 | 0.390 | 0.309 | 0.244 | 0.290 | 0.350 | 0.317 | 0.317 | 0.300 |
| 1985 | 0.044 | 0.433 | 0.391 | 0.294 | 0.352 | 0.342 | 0.212 | 0.277 | 0.232 | 0.284 | 0.284 | 0.329 |
| 1986 | 0.071 | 0.330 | 0.419 | 0.348 | 0.300 | 0.312 | 0.238 | 0.226 | 0.370 | 0.290 | 0.290 | 0.310 |
| 1987 | 0.092 | 0.312 | 0.345 | 0.292 | 0.253 | 0.314 | 0.322 | 0.182 | 0.205 | 0.256 | 0.256 | 0.288 |
| 1988 | 0.128 | 0.376 | 0.354 | 0.362 | 0.398 | 0.273 | 0.349 | 0.224 | 0.201 | 0.289 | 0.289 | 0.333 |
| 1989 | 0.116 | 0.352 | 0.565 | 0.496 | 0.411 | 0.393 | 0.286 | 0.363 | 0.337 | 0.359 | 0.359 | 0.409 |
| 1990 | 0.136 | 0.379 | 0.373 | 0.473 | 0.428 | 0.346 | 0.410 | 0.417 | 0.413 | 0.404 | 0.404 | 0.404 |
| 1991 | 0.068 | 0.218 | 0.318 | 0.285 | 0.309 | 0.256 | 0.212 | 0.311 | 0.230 | 0.264 | 0.264 | 0.273 |
| 1992 | 0.089 | 0.286 | 0.316 | 0.293 | 0.156 | 0.229 | 0.237 | 0.191 | 0.167 | 0.196 | 0.196 | 0.244 |
| 1993 | 0.068 | 0.260 | 0.333 | 0.423 | 0.326 | 0.283 | 0.221 | 0.341 | 0.194 | 0.274 | 0.274 | 0.312 |
| 1994 | 0.044 | 0.279 | 0.302 | 0.248 | 0.206 | 0.230 | 0.126 | 0.223 | 0.405 | 0.238 | 0.238 | 0.231 |
| 1995 | 0.029 | 0.171 | 0.418 | 0.353 | 0.261 | 0.349 | 0.317 | 0.202 | 0.171 | 0.261 | 0.261 | 0.296 |
| 1996 | 0.102 | 0.167 | 0.279 | 0.316 | 0.298 | 0.231 | 0.275 | 0.288 | 0.221 | 0.263 | 0.263 | 0.265 |


| year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | $\mathbf{F}_{\text {bar }}$ <br> (3-9) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0.071 | 0.317 | 0.397 | 0.383 | 0.285 | 0.382 | 0.214 | 0.254 | 0.132 | 0.254 | 0.254 | 0.319 |
| 1998 | 0.065 | 0.257 | 0.309 | 0.288 | 0.403 | 0.366 | 0.237 | 0.216 | 0.295 | 0.304 | 0.304 | 0.297 |
| 1999 | 0.086 | 0.298 | 0.311 | 0.352 | 0.329 | 0.357 | 0.403 | 0.223 | 0.255 | 0.314 | 0.314 | 0.325 |
| 2000 | 0.051 | 0.238 | 0.379 | 0.352 | 0.313 | 0.200 | 0.247 | 0.449 | 0.302 | 0.303 | 0.303 | 0.311 |
| 2001 | 0.029 | 0.304 | 0.305 | 0.411 | 0.414 | 0.378 | 0.259 | 0.391 | 0.507 | 0.391 | 0.391 | 0.352 |
| 2002 | 0.096 | 0.323 | 0.270 | 0.338 | 0.453 | 0.319 | 0.358 | 0.364 | 0.375 | 0.375 | 0.375 | 0.346 |
| 2003 | 0.124 | 0.328 | 0.387 | 0.304 | 0.162 | 0.158 | 0.175 | 0.304 | 0.396 | 0.240 | 0.240 | 0.260 |
| 2004 | 0.157 | 0.307 | 0.262 | 0.279 | 0.386 | 0.282 | 0.350 | 0.298 | 0.459 | 0.356 | 0.356 | 0.309 |
| 2005 | 0.070 | 0.251 | 0.366 | 0.404 | 0.346 | 0.343 | 0.285 | 0.384 | 0.332 | 0.339 | 0.339 | 0.340 |
| 2006 | 0.121 | 0.281 | 0.382 | 0.438 | 0.366 | 0.327 | 0.335 | 0.320 | 0.447 | 0.360 | 0.360 | 0.350 |
| 2007 | 0.055 | 0.277 | 0.423 | 0.384 | 0.450 | 0.428 | 0.286 | 0.304 | 0.317 | 0.353 | 0.353 | 0.365 |
| 2008 | 0.073 | 0.268 | 0.330 | 0.342 | 0.340 | 0.385 | 0.425 | 0.224 | 0.275 | 0.347 | 0.347 | 0.330 |
| 2009 | 0.047 | 0.175 | 0.191 | 0.239 | 0.278 | 0.200 | 0.235 | 0.233 | 0.216 | 0.202 | 0.202 | 0.222 |
| 2010 | 0.015 | 0.119 | 0.167 | 0.221 | 0.202 | 0.252 | 0.201 | 0.327 | 0.173 | 0.198 | 0.198 | 0.213 |
| 2011 | 0.028 | 0.129 | 0.211 | 0.231 | 0.215 | 0.210 | 0.222 | 0.218 | 0.232 | 0.340 | 0.340 | 0.205 |
| 2012 | 0.009 | 0.081 | 0.175 | 0.247 | 0.291 | 0.305 | 0.256 | 0.263 | 0.359 | 0.392 | 0.392 | 0.231 |
| 2013 | 0.039 | 0.113 | 0.222 | 0.273 | 0.244 | 0.255 | 0.289 | 0.261 | 0.273 | 0.284 | 0.284 | 0.237 |
| 2014 | 0.062 | 0.125 | 0.212 | 0.278 | 0.273 | 0.203 | 0.184 | 0.221 | 0.231 | 0.216 | 0.216 | 0.214 |
| 2015 | 0.046 | 0.125 | 0.201 | 0.209 | 0.244 | 0.191 | 0.186 | 0.170 | 0.168 | 0.153 | 0.153 | 0.189 |
| 2016 | 0.017 | 0.131 | 0.235 | 0.189 | 0.211 | 0.333 | 0.241 | 0.140 | 0.140 | 0.163 | 0.163 | 0.211 |
| 2017 | 0.027 | 0.276 | 0.131 | 0.213 | 0.261 | 0.145 | 0.187 | 0.068 | 0.094 | 0.559 | 0.559 | 0.183 |
| 2018 | 0.009 | 0.115 | 0.129 | 0.135 | 0.342 | 0.207 | 0.388 | 0.271 | 0.167 | 1.108 | 1.108 | 0.227 |
| 2019 | 0.019 | 0.119 | 0.221 | 0.269 | 0.253 | 0.306 | 0.318 | 0.309 | 0.279 | 0.237 | 0.237 | 0.257 |
| 2020 | 0.020 | 0.109 | 0.212 | 0.294 | 0.273 | 0.239 | 0.319 | 0.229 | 0.267 | 0.184 | 0.184 | 0.239 |
| 2021 | 0.045 | 0.117 | 0.173 | 0.299 | 0.438 | 0.297 | 0.227 | 0.462 | 0.322 | 0.257 | 0.257 | 0.287 |

Table 32.10. Sole in Division 7.e. Assessment summary.

| Year | Recruitment Age 2 [thousands] | TSB [tonnes] | SSB [tonnes] | Landings [tonnes] | Yield/SSB | $\begin{gathered} \text { Fbar } \\ \text { (Ages 3-9) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 1874 | 2927 | 2437 | 353 | 0.14 | 0.134 |
| 1970 | 1343 | 3023 | 2652 | 391 | 0.15 | 0.146 |
| 1971 | 3826 | 2838 | 2390 | 432 | 0.18 | 0.181 |
| 1972 | 2568 | 3091 | 2395 | 437 | 0.18 | 0.138 |
| 1973 | 2264 | 3266 | 2778 | 459 | 0.17 | 0.160 |
| 1974 | 3107 | 3512 | 2896 | 427 | 0.15 | 0.153 |
| 1975 | 2967 | 4428 | 3670 | 491 | 0.13 | 0.131 |
| 1976 | 2791 | 4102 | 3403 | 616 | 0.18 | 0.180 |
| 1977 | 6556 | 5339 | 4098 | 606 | 0.15 | 0.130 |
| 1978 | 4657 | 5428 | 4074 | 861 | 0.21 | 0.178 |
| 1979 | 4389 | 6014 | 4865 | 1181 | 0.24 | 0.242 |
| 1980 | 4702 | 6387 | 5338 | 1269 | 0.24 | 0.232 |
| 1981 | 8130 | 5957 | 4572 | 1215 | 0.27 | 0.270 |
| 1982 | 4679 | 5916 | 4575 | 1446 | 0.32 | 0.321 |
| 1983 | 3866 | 5377 | 4374 | 1498 | 0.34 | 0.363 |
| 1984 | 5968 | 5462 | 4430 | 1370 | 0.31 | 0.300 |
| 1985 | 6982 | 5568 | 4009 | 1409 | 0.35 | 0.329 |
| 1986 | 3765 | 5257 | 4013 | 1419 | 0.35 | 0.310 |
| 1987 | 5847 | 5310 | 4111 | 1280 | 0.31 | 0.288 |
| 1988 | 3878 | 5119 | 4043 | 1444 | 0.36 | 0.333 |
| 1989 | 3735 | 4318 | 3442 | 1390 | 0.4 | 0.409 |
| 1990 | 2817 | 4222 | 3287 | 1315 | 0.4 | 0.404 |
| 1991 | 7161 | 4219 | 2991 | 852 | 0.28 | 0.273 |
| 1992 | 3902 | 4100 | 2937 | 895 | 0.3 | 0.244 |
| 1993 | 3350 | 3579 | 2810 | 904 | 0.32 | 0.312 |
| 1994 | 2378 | 3785 | 3053 | 800 | 0.26 | 0.231 |
| 1995 | 3452 | 3875 | 3067 | 856 | 0.28 | 0.296 |
| 1996 | 3940 | 4152 | 3054 | 833 | 0.27 | 0.265 |


| Year | Recruitment Age 2 [thousands] | TSB [tonnes] | SSB [tonnes] | Landings [tonnes] | Yield/SSB | $\begin{gathered} \text { Fbar } \\ \text { (Ages 3-9) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 3331 | 3831 | 2920 | 949 | 0.32 | 0.319 |
| 1998 | 4398 | 3941 | 2910 | 880 | 0.3 | 0.297 |
| 1999 | 3585 | 3947 | 2830 | 957 | 0.34 | 0.325 |
| 2000 | 6522 | 4301 | 2865 | 914 | 0.32 | 0.311 |
| 2001 | 5423 | 4507 | 2894 | 1069 | 0.37 | 0.352 |
| 2002 | 3815 | 4183 | 3009 | 1106 | 0.37 | 0.346 |
| 2003 | 5381 | 4398 | 3287 | 1078 | 0.33 | 0.260 |
| 2004 | 2870 | 4021 | 3099 | 1075 | 0.35 | 0.309 |
| 2005 | 3994 | 3963 | 3087 | 1039 | 0.34 | 0.340 |
| 2006 | 4612 | 3688 | 2717 | 1023 | 0.38 | 0.350 |
| 2007 | 3934 | 3779 | 2751 | 1015 | 0.37 | 0.365 |
| 2008 | 4209 | 3801 | 2679 | 908 | 0.34 | 0.330 |
| 2009 | 3788 | 3920 | 2967 | 701 | 0.24 | 0.222 |
| 2010 | 4864 | 4433 | 3430 | 698 | 0.2 | 0.213 |
| 2011 | 3471 | 4716 | 3555 | 801 | 0.23 | 0.205 |
| 2012 | 3540 | 4570 | 3738 | 872 | 0.23 | 0.231 |
| 2013 | 3282 | 4513 | 3662 | 883 | 0.24 | 0.237 |
| 2014 | 3464 | 4843 | 3925 | 885 | 0.23 | 0.214 |
| 2015 | 4160 | 4955 | 3988 | 774 | 0.19 | 0.189 |
| 2016 | 5817 | 6005 | 4523 | 913 | 0.2 | 0.211 |
| 2017 | 4908 | 5958 | 4293 | 1007 | 0.23 | 0.183 |
| 2018 | 9132 | 6671 | 4605 | 1075 | 0.23 | 0.227 |
| 2019 | 6378 | 6838 | 4946 | 1185 | 0.24 | 0.257 |
| 2020 | 9050 | 7626 | 5464 | 1219 | 0.22 | 0.239 |
| 2021 | 2745 | 6702 | 5418 | 1392 | 0.26 | 0.287 |

Table 32.11. Sole in Division 7.e. Input data for the short-term forecast.

| Age | N2022 | N2023 | N2024 | M | Mat | PF | PM | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 4077 | 4077 | 4077 | 0.1 | 0.14 | 0 | 0 | 0.148 | 0.031 | 0.175 |
| 3 | 2374 | 3577 | 3576 | 0.1 | 0.45 | 0 | 0 | 0.202 | 0.127 | 0.227 |
| 4 | 6464 | 1893 | 2848 | 0.1 | 0.88 | 0 | 0 | 0.252 | 0.222 | 0.277 |
| 5 | 3497 | 4683 | 1368 | 0.1 | 0.98 | 0 | 0 | 0.3 | 0.316 | 0.323 |
| 6 | 3232 | 2306 | 3080 | 0.1 | 1 | 0 | 0 | 0.345 | 0.354 | 0.368 |
| 7 | 996 | 2053 | 1460 | 0.1 | 1 | 0 | 0 | 0.389 | 0.309 | 0.41 |
| 8 | 905 | 661 | 1360 | 0.1 | 1 | 0 | 0 | 0.429 | 0.317 | 0.448 |
| 9 | 646 | 597 | 435 | 0.1 | 1 | 0 | 0 | 0.467 | 0.367 | 0.485 |
| 10 | 198 | 405 | 373 | 0.1 | 1 | 0 | 0 | 0.502 | 0.319 | 0.519 |
| 11 | 201 | 130 | 266 | 0.1 | 1 | 0 | 0 | 0.535 | 0.249 | 0.55 |
| 12 | 733 | 659 | 556 | 0.1 | 1 | 0 | 0 | 0.63 | 0.249 | 0.639 |

Table 32.12. Sole in Division 7.e. Single option output of the short-term forecast (targeting $F_{\text {msy }}$ ).

| Age | $F$ | Catch.No | Yield | Stock.No | Biomass | SSNo | SSB |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Year $=2022, F_{\text {bar }}=0.287$

| 2 | 0.031 | 118 | 21 | 4077 | 605 | 571 | 85 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.127 | 269 | 61 | 2374 | 479 | 1068 | 215 |
| 4 | 0.222 | 1228 | 340 | 6464 | 1629 | 5688 | 1433 |
| 5 | 0.316 | 904 | 292 | 3497 | 1050 | 3427 | 1029 |
| 6 | 0.354 | 919 | 338 | 3232 | 1116 | 3232 | 1116 |
| 7 | 0.309 | 253 | 103 | 996 | 387 | 996 | 387 |
| 8 | 0.317 | 235 | 105 | 905 | 388 | 905 | 388 |
| 9 | 0.367 | 190 | 92 | 646 | 302 | 646 | 302 |
| 10 | 0.319 | 51 | 27 | 198 | 99 | 198 | 99 |
| 11 | 0.249 | 42 | 23 | 201 | 108 | 201 | 108 |
| 12 | 0.249 | 154 | 98 | 733 | 462 | 733 | 462 |
| Total | NA | 4363 | 1500 | 23323 | 6624 | 17665 | 5624 |

Year $=2023, F_{\text {bar }}=0.290$

| 2 | 0.031 | 119 | 21 | 4077 | 605 | 571 | 85 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.128 | 409 | 93 | 3577 | 721 | 1610 | 325 |
| 4 | 0.224 | 363 | 100 | 1893 | 477 | 1665 | 420 |
| 5 | 0.319 | 1220 | 394 | 4683 | 1406 | 4589 | 1378 |
| 6 | 0.357 | 661 | 243 | 2306 | 797 | 2306 | 797 |
| 7 | 0.312 | 525 | 215 | 2053 | 798 | 2053 | 798 |
| 8 | 0.32 | 173 | 77 | 661 | 284 | 661 | 284 |
| 9 | 0.371 | 176 | 86 | 597 | 279 | 597 | 279 |
| 10 | 0.322 | 106 | 55 | 405 | 203 | 405 | 203 |
| 11 | 0.251 | 27 | 15 | 130 | 70 | 130 | 70 |
| 12 | 0.251 | 139 | 89 | 659 | 415 | 659 | 415 |
| Total | NA | 3918 | 1388 | 21042 | 6054 | 15247 | 5052 |


| Age | F | Catch.No | Yield | Stock.No | Biomass | SSNo | SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year $=2024, \mathrm{~F}_{\text {bar }}=0.290$ |  |  |  |  |  |  |  |
| 2 | 0.031 | 119 | 21 | 4077 | 605 | 571 | 85 |
| 3 | 0.128 | 409 | 93 | 3576 | 721 | 1609 | 325 |
| 4 | 0.224 | 546 | 151 | 2848 | 718 | 2507 | 632 |
| 5 | 0.319 | 357 | 115 | 1368 | 411 | 1341 | 403 |
| 6 | 0.357 | 883 | 324 | 3080 | 1064 | 3080 | 1064 |
| 7 | 0.312 | 373 | 153 | 1460 | 568 | 1460 | 568 |
| 8 | 0.32 | 355 | 159 | 1360 | 583 | 1360 | 583 |
| 9 | 0.371 | 129 | 62 | 435 | 203 | 435 | 203 |
| 10 | 0.322 | 98 | 51 | 373 | 187 | 373 | 187 |
| 11 | 0.251 | 56 | 31 | 266 | 142 | 266 | 142 |
| 12 | 0.251 | 118 | 75 | 556 | 350 | 556 | 350 |
| Total | NA | 3441 | 1235 | 19400 | 5552 | 13558 | 4541 |

Units are thousands (for numbers) and tonnes (for weights).

Table 32.13. Sole in Division 7.e. Year-class sources and contributions for the short-term forecast (in percent).

| cohort | Yield 2022 | Yield 2023 | SSB 2022 | SSB 2023 | SSB 2024 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 6.5 | NA | 8.2 | NA | NA |
| 2011 | 1.5 | 6.4 | 1.9 | 8.2 | NA |
| 2012 | 1.8 | 1.1 | 1.8 | 1.4 | 7.7 |
| 2013 | 6.1 | 4 | 5.4 | 4 | 3.1 |
| 2014 | 7 | 6.2 | 6.9 | 5.5 | 4.1 |
| 2015 | 6.9 | 5.6 | 6.9 | 5.6 | 4.5 |
| 2016 | 22.5 | 15.5 | 19.8 | 15.8 | 12.8 |
| 2017 | 19.5 | 17.5 | 18.3 | 15.8 | 12.5 |
| 2018 | 22.6 | 28.4 | 25.5 | 27.3 | 23.4 |
| 2019 | 4.1 | 7.2 | 3.8 | 8.3 | 8.9 |
| 2020 | 1.4 | 6.7 | 1.5 | 6.4 | 13.9 |
| 2021 |  | 1.5 |  | 1.7 | 7.1 |
| 2022 |  |  |  |  | 1.9 |



Table 32.14. Sole in Division 7.e. Annual catch scenarios. All weights are in tonnes.

| Basis | Total catch* (2023) | Pro- <br> jected <br> landings (2023) | Projected discards (2023) | $F_{\text {projected }}$ landings (2023) | $\begin{aligned} & \text { SSB } \\ & \text { (2024) } \end{aligned}$ | \% SSB <br> change <br> ** | \% TAC change*** | \% advice change^ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |
| MSY approach: $\mathrm{F}_{\text {MSY }}$ | 1394 | 1388 | 6 | 0.29 | 4541 | -10.1 | -23 | -23 |
| Other options |  |  |  |  |  |  |  |  |
| EU MAP^^: $\mathrm{F}_{\text {MSY }}$ | 1394 | 1388 | 6 | 0.29 | 4541 | -10.1 | -23 | -23 |
| $F=M A P \wedge \wedge ~ F_{\text {MSY lower }}$ | 820 | 816 | 4 | 0.160 | 5103 | 1.01 | -55 | -55 |
| $F=M A P \wedge \wedge ~ F_{\text {MSY upper }}$ | 1596 | 1589 | 7 | 0.34 | 4344 | -14.0 | -11.8 | -11.8 |
| $F=0$ | 0 | 0 | 0 | 0 | 5909 | 17.0 | -100 | -100 |
| $\mathrm{F}_{\mathrm{pa}}$ | 1788 | 1781 | 8 | 0.39 | 4156 | -17.7 | -1.19 | -1.19 |
| Flim | 1971 | 1962 | 9 | 0.44 | 3979 | -21 | 8.9 | 8.9 |
| $\mathrm{SSB}_{2024}=\mathrm{Bl}_{\text {lim }}$ | 4036 | 4018 | 18 | 1.29 | 2000 | -60 | 123 | 123 |
| Rollover TAC | 1810 | 1802 | 8 | 0.40 | 4135 | -18.1 | 0 | 0 |
| $\begin{aligned} & \operatorname{SSB}_{2024}=\mathrm{B}_{\mathrm{pa}}=\text { gSY }_{\text {ger }} \end{aligned}$ | 3088 | 3074 | 14 | 0.81 | 2900 | -43 | 71 | 71 |
| $S S S B 2024=$ SSB $_{2023}$ | 872 | 868 | 4 | 0.171 | 5052 | 0 | -52 | -52 |

* Total catch derived from the projected landings and the assumed discard rate.
** SSB 2024 relative to SSB 2023.
*** Total catch in 2023 relative to TAC 2022 (1810 tonnes).
^ Advice value for 2023 relative to the advice value for 2022 ( 1810 tonnes).
${ }^{\wedge \wedge}$ EU multiannual plan (MAP) for the Western Waters (EU, 2019).

Table 32.15. Sole in Division 7.e. Annual catch scenarios (more options and more digits provided, sorted by fishing mortality in intermediate year). All weights are in tonnes.

| Basis | Total catch* (2023) |
| :--- | ---: | :--- | ---: | :--- |
| Projected landings |  |
| (2023) |  |


| Basis | Total catch* (2023) | Projected landings (2023) | Projected discards (2023) | $F_{\text {projected landings }}$ (2023) | SSB (2024) | \% SSB change ** | \% TAC change*** | \% advice change^ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{F}_{\text {sq }}$ | 1383.792056 | 1377.724 | 6.068163 | 0.28743 | 4550.977 | -9.91775 | -23.5474 | -23.5474 |
| $\mathrm{F}_{\text {MSY }}$ | 1394.443516 | 1388.329 | 6.114872 | 0.29 | 4540.573 | -10.1237 | -22.9589 | -22.9589 |
| $\mathrm{F}=0.3$ | 1435.633468 | 1429.338 | 6.295496 | 0.3 | 4500.348 | -10.9199 | -20.6832 | -20.6832 |
| $\mathrm{F}=0.31$ | 1476.413142 | 1469.939 | 6.474322 | 0.31 | 4460.537 | -11.7079 | -18.4302 | -18.4302 |
| $\mathrm{F}=0.32$ | 1516.786901 | 1510.136 | 6.651368 | 0.32 | 4421.134 | -12.4879 | -16.1996 | -16.1996 |
| TAC085 | 1538.5 | 1531.753 | 6.746584 | 0.32542 | 4399.948 | -12.9072 | -15 | -15 |
| $\mathrm{F}=0.33$ | 1556.75906 | 1549.932 | 6.826653 | 0.33 | 4382.135 | -13.2598 | -13.9912 | -13.9912 |
| $\mathrm{F}_{\text {MSY upper }}$ | 1596.333886 | 1589.334 | 7.000195 | 0.34 | 4343.537 | -14.0238 | -11.8048 | -22.8451 |
| Fpa | 1788.39415 | 1780.552 | 7.842412 | 0.39 | 4156.394 | -17.7281 | -1.19369 | -1.19369 |
| TAC | 1810 | 1802.063 | 7.937157 | 0.395784 | 4135.36 | -18.1445 | -1.11E-14 | -1.11E-14 |
| Flim | 1971.134969 | 1962.491 | 8.643761 | 0.44 | 3978.624 | -21.2469 | 8.902484 | 8.902484 |
| TAC115 | 2081.5 | 2072.372 | 9.127731 | 0.471443 | 3871.409 | -23.3691 | 15 | 15 |
| $\mathrm{B}_{\mathrm{pa}}$ | 3087.903683 | 3074.363 | 13.54098 | 0.814384 | 2900 | -42.5973 | 70.60241 | 70.60241 |
| $\mathrm{B}_{\text {trigger }}$ | 3087.903683 | 3074.363 | 13.54098 | 0.814384 | 2900 | -42.5973 | 70.60241 | 70.60241 |
| Blim | 4036.128381 | 4018.429 | 17.69911 | 1.29102 | 2000 | -60.4119 | 122.9905 | 122.9905 |

* Total catch derived from the projected landings and the assumed discard rate.
** SSB 2024 relative to SSB 2023.
*** Total catch in 2023 relative to TAC 2022 ( 1810 tonnes).
${ }^{\wedge}$ Advice value for 2023 relative to the advice value for 2022 ( $\mathbf{1 8 1 0}$ tonnes).


### 32.16 Figures



Figure 32.1. Sole in Division 7.e. Landings and discards reported in InterCatch by country.


Figure 32.2. Sole in Division 7.e. International landings reported in InterCatch by fleet and year.


Figure 32.3. Sole in Division 7.e. Discard rates for discards reported in InterCatch.


Figure 32.4. Sole in Division 7.e. Annual reported discard rates in InterCatch by fleet and country.


Figure 32.5. Sole in Division 7.e. International landings numbers-at-age (last 16 years).


Figure 32.6. Sole in Division 7.e. Catch (landings) and stock weights-at-age.


Figure 32.7. Sole in Division 7.e. Generation of stock and catch weights from landings weights-at-age.


Figure 32.8. Sole in Division 7.e. Landings age distributions submitted to InterCatch. Numbers are raised to fleet level.


Figure 32.9. Sole in Division 7.e. Length distributions submitted to InterCatch. Numbers are raised to fleet level.


Figure 32.10. Sole in Division 7.e. Means standardised Ipue and effort for the UK commercial fleets.

Figure 32.11. Sole in Division 7.e. Means standardised lpue/cpue by year class. Note, the cohorts differ on the x-axes due to the differences in the length and age ranges of the tuning series.


age

- 11
- 10
- 9
- 8
- 7
- 6
- 5
- 4
- 3
- 2



Figure 32.12. Sole in Division 7.e. Means standardised Ipue/cpue by year. Note, the lines differ on the x-axes due to the differences in the length and age ranges of the tuning series.


Figure 32.13. Sole in Division 7.e. Survey indices (raw values) for all commercial and scientific surveys. The plots on the left show the index values at-age, on the right are the values aggregated over all ages.


Figure 32.14. Sole in Division 7.e. Internal consistencies in the scientific surveys. Shown is the correlation between num-bers-at-age and the numbers of the same cohort one year later, including Pearson correlation coefficient $\rho$ and the $p$ value.


Figure 32.15. Sole in Division 7.e. Internal consistencies in the scientific surveys. Shown is the correlation between num-bers-at-age and the numbers of the same cohort one year later, including Pearson correlation coefficient $\rho$ and the p value.


Figure 32.16. Sole in Division 7.e. Internal consistencies in the commercial surveys. Shown is the correlation between numbers-at-age and the numbers of the same cohort one year later, including Pearson correlation coefficient $\rho$ and the $p$-value.


Figure 32.17. Sole in Division 7.e. Sole catch rates during FSP "Western Channel Sole and Plaice" surveys, 2003-2021 (number $\mathrm{h}^{-1} \mathrm{~m}^{\text {beam }}{ }^{-1}$ ). Open circles: FV Nellie and FV Carhelmar tows; filled circles: FV Lady T Emiel tows. Please note that 2021 numbers are not to scale. Source: Burt et al. (2021, 2022).


Figure 32.18. Sole in Division 7.e. Results of the final XSA run. Summary of the stock assessment. ICES estimated catches, recruitment (age 2), fishing mortality (F), and spawning-stock biomass (SSB). The assumed recruitment value for 2022 is shaded in a lighter colour. Discard estimates are only available since 2012.




$\log$ residuals

- 0.01
- 0.10

○ 0.50
O 1.00
○ 2.00

- negative
- positive
age
- 11
$-\quad 10$
$---\quad 9$
-+-8
… 7
$\rightarrow-6$
-- 5
$-\quad 4$
$-\quad 3$
$-\quad$.
-. 2

Figure 32.19. Sole in Division 7.e. XSA fleet log catchability residuals for. Note that the application of time-series weighting set as a tricubic taper with a range of 15 years excludes log catchability residuals prior to 2004.

## SSB (1000 t)



F (ages 3-9)


Rec (age 2; Millions)


Figure 32.20. Sole in Division 7.e. Comparison of the current XSA assessment with the final assessment runs from the last years.


Figure 32.21. Sole in Division 7.e. Scaled weights for the current XSA assessment and the previous XSA assessment conducted at last year's WGCSE.


Figure 32.22. Sole in Division 7.e. Five-year retrospective of stock status and fishing mortality estimates.


Figure 32.23. Sole in Division 7.e. Options for the intermediate year in the short-term forecast.


Figure 32.24. Sole in Division 7.e. Comparison of international TAC, catch advice and realised landings.


Figure 32.25. Sole in Division 7.e. Output for the short-term forecast under the MSY approach.


Figure 32.26. Sole in Division 7.e. Output of the short-term forecast of the MSY approach, including F $_{\text {MSY }}$ ranges.


Figure 32.27. Sole in Division 7.e. Age class contributions (biomass) to the SSB and catch. Age $\mathbf{1 2}$ is the plus group. The vertical black line indicates where the short-term forecast starts.


Figure 32.28. Sole in Division 7.e. Age class contributions (numbers) to the stock and catch. Age $\mathbf{1 2}$ is the plus group. The vertical black line indicates where the short-term forecast starts.

## 33 Sole (Solea solea) in Division 7.f and 7.g (Bristol Channel, Celtic Sea)

## Type of assessment in 2022

This assessment is an update assessment.
ICES advice applicable to 2022
ICES advises that when the MSY approach is applied, catches in 2022 should be no more than $1337^{1}$ tonnes.

## Comments made by the audit of last year's assessment

No major deficiencies for the sole assessment in the Celtic sea were reported.

### 33.1 General

Stock description and management units.


The sole fisheries in the Celtic Sea are managed by TAC and technical measures. A TAC is in place for ICES divisions 7.f and 7.g. These divisions do correspond to the stock area. The basis for the stock assessment area 7.f and 7.g is described in detail in the Stock Annex.
The agreed TACs in 2021 and 2022 are presented in the text tables below. Technical measures in force for this stock are minimum mesh sizes and minimum conservation reference sizes (MCRS, 25 cm for Belgian vessels from March 11th 2017 onwards, except vessels with engine power $<221 \mathrm{~kW}$ and/or volume $<70 \mathrm{GT}$ ). National regulations also restricted areas for certain types of vessels.

Three rectangles in the Celtic Sea (30E4, 31E4 and 32E3, referred to as the "Trevose Box") were closed during the first quarter of 2005, and in February-March each year from 2006 onwards. A derogation has permitted beam trawlers to fish there in March 2005. The effects of this closure have been discussed in previous WGSSDS meetings and ACFM 2007, and evaluated at WKCELT 2014

[^18]
## Management applicable to 2021 and 2022

The TAC and the national quotas by country for 2021.

| Species: | Common sole <br> Solea solea |  | Zone: <br> 7f and 7g <br> (SOL/7FG.) |
| :--- | :--- | :--- | :--- |
| Belgium | 830 | Analytical TAC |  |
| France | 83 | Article 8(2) of this Regulation applies |  |
| Ireland | 42 |  |  |
| Union | 955 |  |  |
| United Kingdom | 433 |  |  |
|  |  |  |  |
| TAC |  |  |  |

The TAC and the national quotas by country for 2022.

| Species: | Common sole <br> Solea solea | Zone:7f and 7g <br> (SOL/7FG.) |
| :--- | :---: | :--- |
| Belgium | 781 | Analytical TAC <br> France |
| Article 8(2) of this Regulation applies |  |  |
| Ireland | 78 |  |
| Union | 898 |  |
| United Kingdom | 415 |  |
| TAC | 1337 |  |

## Fishery in 2021

An overview of the landings and discard data provided and used by the Working Group (WG) is shown in Table 33.1 and Figure 33.1. The landings have fluctuated over the time-series with higher amounts of around 1500-1600 t in 1986, 2003 and 2020. The available discard data indicate that discarding of sole has increased in 2018 and 2019 (to 141 and 145 t respectively) due to the strong 2016 year class. Afterwards discards decreased again to 106 t in 2020.

In 2021, the WG estimated landings are 1336 t , of which Belgium landed 62.1\% (830 t), UK (England and Wales) $29.6 \%$ ( 395 t ), France $4.5 \%$ ( 60 t ), Ireland $3.6 \% ~(48 \mathrm{t}$ ) and the remainder by Northern Ireland and Spain. Discards were estimated to be at 62 t . This catch figure ( 1398 t ) corresponds to an international uptake of $99 \%$ of the agreed TAC in 2021 (1413 t).

The WG estimate of the 2020 landings and discards was not revised.
In 2021, $87 \%$ of the landings and discards were taken by beam trawls, $13 \%$ by otter trawls and $<1 \%$ by other gears.

### 33.2 Data

## Age-compositions and weights-at-age

InterCatch was used for estimation of both landings and discards numbers and age compositions, as input for the assessment. Belgium, Ireland, France, UK, Spain and Northern Ireland have provided data this year under the ICES InterCatch format on a métier basis. Quarterly/yearly data for 2021 were available for landing numbers and weight-at-age, for most of the Belgian, Irish and UK fleets. These comprise $90 \%$ of the international landings. Discard weights were available for $60 \%$ of the landings. The age coverage for the sampled discards is $100 \%$.

If discards were not included for a particular year-quarter-country-métier combination, they are assumed to be unknown (non-zero) and therefore raised (InterCatch). The weighting factor for raising the discards was 'Landings CATON' (landings catch). Discard raising was performed on a gear level regardless of season or country. The following groups were distinguished based on gear:

- TBB
- OTB including OTB, OTT, SSC, SDN
- GTR including GTR and GNS

The remaining gears were combined in a REST group (including MIS, FPO, DRB, LHM, LLS).
The GNS/GTR, TBB and OTB/OTT/SSC/SDN groups contribute respectively $0 \%, 69 \%$ and $0.006 \%$ to the landings of sole in 27.7.fg.

Raising within a gear group was performed when the proportion of landings for which discard weights are available was equal or larger than $50 \%$ compared to the total landings of that group. For the 2021 data, this was only the case for the TBB gear group. When the threshold was not reached for a gear group, it was pooled with the REST group to raise discards based on all available information.

To allocate age compositions, landings and discards were handled separately; samples from landings were used only for landings and vice versa. When age distributions (both landings and discards) had to be borrowed from other strata, allocations were performed on a gear level. The same gear groups (TBB, OTB, GTR and REST) as used for discard raising were applied. In 2021, the proportion of landings covered for age composition is respectively $27 \%, 100 \%$ and $22 \%$ for the GNS/GTR, TBB and OTB/OTT/SSC/SDN group. The proportion of landings for which discard age coverage is available is respectively $0 \%, 69 \%$, and $0 \%$ for the GNS/GTR, TBB and OTB/OTT/SSC/SDN group. When the threshold of $50 \%$ was reached for the proportion of landings or discards covered by age, allocation of age occurred with all available information within that gear group. For the 2021 landings and discard data, this threshold was only reached for the TBB group. When the threshold was not reached, unsampled data were pooled in the REST group and ages were allocated using all sampled data. The weighting factor was 'Mean Weight weighted by numbers-at-age'.

Figure. 33.2 shows the available landings and discards data by country, gear and year.
Raised discard data from InterCatch were available from 2004 onwards. To estimate discard mean weight-at-age and numbers-at-age prior to 2004, a constant ratio of discards to landings by age was applied using data from 2004-2018 (WKFlatNSCS 2020).

Further details on raising methods are given in the stock annex.

Catch numbers-at-age are given in Table 33.2 and weights-at-age in the catch are given in Table 33.3. Age compositions are plotted in Figures 33.3ab. The standardised catch proportion-at-age is presented in Figure 33.4.

## Length-compositions

Annual length compositions for 2021 are given by fleet in Table 33.4 Length distributions of the total Belgian and UK (E\&W) landings for the last 21 years are plotted in Figure 33.5. Belgian vessels generally land a greater proportion of small fish compared to the UK(England and Wales).

The length distributions for 2021 of retained and discarded catches of sole by the Belgium beam trawl fleet are presented in Figure 33.6 from The Belgian beam trawl fleet mainly discarded fish of 23 and 24 cm . According to the Belgian age-length samples, these fish were mainly age 3 or 4 .

## Biological

The stock weights (Table 33.5) were obtained using the Rivard weight calculator (http://nft.nefsc.noaa.gov./), that conducts a cohort interpolation of the catch weights. The resulting stock weight for age 1 was very variable, and it was decided during the benchmark to set the stock weight of age 1 to the lowest estimated stock weight for age 2 for 1971-2019.

A new maturity ogive was estimated during the WKFlatNSCS (ICES, 2020) using only survey data of the UK(E\&W)-Q1SWECOS. Maturity data are available for 2013-2019. The new maturity ogive is calculated with a length-based model with sex-specific ALK. This new ogive indicates that $>60 \%$ of the 2 and 3 year old individuals are mature, while this was not the case in the maturity ogive used until the WGCSE 2019. The maturity at-age 1 was manually set to 0 as no mature sole at age 1 were encountered at the UK(E\&W)-Q1SWECOS survey.

Updated maturity at-age based on data from the UK(E\&W)-Q1SWECOS survey.

| Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Maturity | 0.0 | 0.67 | 0.91 | 0.98 | 0.99 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Natural mortality was assumed to be 0.1 for all ages and years

## Surveys

The WGCSE 2022 Celtic Sea sole stock assessment used one scientific survey index: UK(E\&W)-BTS-Q3 (1988-2021), from age 1 to 5. It is the only index providing information on the recruiting age (age 1). Standardised abundance indices for the UK beam trawl survey (UK(E\&W)-BTS-Q3)) are shown in Table 33.6 and Figure 33.7. Abundance-at-age 0 is highly variable and not used further on. The UK-survey appears to track the stronger year classes reasonably well. The internal consistency plot indicates also a reasonable fit for most of the ages (Figure 33.8).

After the peak in 2000 ( $228.46 \mathrm{~kg} / 100 \mathrm{Km}$ fished, Figure 33.9 and Table 33.8), the LPUE from the UK(E\&W)-BTS-Q3 dropped gradually to the lowest value in $2006(68.967 \mathrm{~kg} / 100 \mathrm{Km}$ fished). Thereafter, it fluctuated between $80 \mathrm{~kg} / 100 \mathrm{Km}$ fished and $120 \mathrm{~kg} / 100 \mathrm{Km}$ fished until 2017. In 2018 , it increased again to $206.44 \mathrm{~kg} / 100 \mathrm{Km}$ fished and for 2020 and 2021 a lower value of about $112 \mathrm{~kg} / 100 \mathrm{Km}$ fished was noted.

Detailed information on the survey protocols and area coverage can be found in the Stock Annex.

## Commercial LPUE

Available estimates of effort and LPUE are presented in Tables 33.7-33.8 and Figure 33.10.
Commercial LPUE and effort data were available for Belgian beam trawlers, UK(E\&W) beam and otter trawlers and Irish seiners, otter and beam trawlers. It should be noted that in 2013, the UK administration switched to the EU electronic logbook system. Therefore, effort and LPUE reporting is now based on days fished.

Belgian beam trawl effort was at highest levels in 2003-2005. During these years, effort shifted from the Eastern English Channel (Division 7.d) to the Celtic Sea (divisions 7.fg) because of days-at-sea limitations in the former area. In 2006, these restrictions had been lifted and effort decreased substantially to about half of the values observed in the early 2000s. The sharp effort reduction in 2008 may be a combined result of the unrestricted effort regime in Division 7.d and the high fuel prices. The increase in 2012-2013 is due to the good opportunities of sole catches in the Celtic Sea taken by the mobile Belgian fleet. Afterwards, effort decreased again to lower levels during 2014-2019. Together with the substantial increase of the TAC in 2020, the effort also increased in 2020. A slightly lower effort was noted in 2021.

The effort from the UK(E\&W) beam trawl fleet has declined sharply since the early 2000s to a record low in 2011 (area 7 f and area 7 g east) and 2008 (area 7 g west), and fluctuated between this lower value and the time-series mean afterwards. For area 7 g a just above average value was noted for 2021. The effort from the UK (E\&W) otter trawlers has shown a gradually declining trend over time.

LPUE of the Belgian beam trawlers peaked in 2002. After a sharp decline to its record low in 2004, LPUE has been increasing gradually to around $20-21 \mathrm{~kg} / \mathrm{hour}$ in 2014-2015. In 2017, a decrease to $15.72 \mathrm{~kg} /$ hour was recorded. Afterwards it increased again and was at the highest level of the time-series in 2020 ( 25.74 kg /hour). A slightly lower value of $22.04 \mathrm{~kg} / \mathrm{hour}$ was recorded in 2021.

At the end of the 1990s and the beginning of the 2000s, the LPUE of the UK beam trawlers was stable at lower levels compared to the period before. Afterwards, the LPUE fluctuated and gradually increased to a value around the time-series mean in 2020/2021.
The LPUE of the UK otter trawlers is relatively stable at a lower level, but increased the last three years in area 7 f and 7 g west.
Irish effort and LPUE data are also presented. The main target species in the Irish fisheries are megrim, anglerfish, etc. The vessels usually operate on fishing grounds in the Western Celtic Sea with lower sole densities and therefore the LPUE values are low.

## Tuning series

All available tuning data are given in Table 33.9, with the data used in the assessment highlighted in bold.

The age-structured UK(E\&W)-BTS-Q3 scientific survey tuning series is the only scientific survey used for tuning.

During the WGCSE 2019, two age-structured commercial tuning series (UK(E\&W)-CBT and BECBT) were used in the assessment. The UK(E\&W)-CBT tuning-series used in the WGCSE 2019 assessment was limited to 2012 and earlier, because of effort reporting issues. As the hours fished became an optional field in the logbooks and not consistently filled, this field is inappropriate to use as a metric for effort.

During the WKFlatNSCS (ICES, 2020) an updated UK(E\&W)-CBT tuning series was introduced in the assessment. The new UK(E\&W)-CBT series from 1987-2021 was generated using a random effects model. Activity days was used as an effort measure, since it is mandatory to record.

The Belgian commercial beam trawl tuning fleet consists of two parts (1971-1996 and 2006-2021, BE_CBT and BE_CBT3). During the IBPBrisol (ICES, 2019b), the BE_CBT3 was constructed focusing on the landings and effort data of pure trips from the large fleet segment of the Belgian beam trawl fleet fishing in divisions 7.f and 7.g. Several models were tested and a GLMM including a categorical year effect, a log-linear relationship between the engine power of a beam trawler and the landing rate, a categorical temporal effect 'month' and a categorical spatial effect 'ICES statistical rectangle' were retained. The exponent of the estimated coefficients of the year effect were used as landing rate for the tuning series. More information is provided in the stock annex and the WKFlatNSCS report (ICES, 2020).
During the Benchmark, these commercial tuning series were used as commercial biomass tuning series. These time-series of the commercial tuning series were split in order to better account for changes in catchability due to e.g. technological creep (see figure below). Figure 33.11 shows the evolution through time of the commercial biomass tuning series. The Belgian BE_CBT_20062021 and the UK(E\&W)-CBT_2006-2021 tuning series show a relatively similar increasing trend during the last years.

## Other relevant data

Reports from UK industry suggest that the main issues affecting the fishery in 7.f and 7.g were displacement of effort due to the rectangle closures and the restrictions on the use of 80 mm mesh west of $7^{\circ} \mathrm{W}$ (Trebilcock and Rozarieux, 2009).

No additional information was received from the Belgian, French and Irish industries.

### 33.3 Stock assessment

## Historical stock development

The method used to assess Celtic Sea sole was XSA until the WGCSE 2019. During the WKFlatNSCS (ICES, 2020), the assessment was transferred to a state-space stock assessment model (SAM). This was done by using the stock assessment package, which enables to interface a performant SAM implementation (https://github.com/fishfollower/SAM/) in Template Model Builder (TMB) ${ }^{2}$ from the R statistical software.

The main feature of SAM is that it includes both process models on survival, recruitment and fishing mortality, describing the internal states of the system, and observation models for catch and tuning data. Additionally, tuning data can be introduced in different ways, e.g. as SSB (spawning-stock biomass), TSB (total stock biomass) or landings indices, while the random effects formulation of the process models resulting from the hierarchical nature of the state-space modelling framework, can easily be used to handle missing observations as is the case with catch information on age 1 . Finally, SAM allows to specify different model configurations, and parametrization of both process and observation models.

[^19]During the benchmark, it was decided to transfer the age-structured commercial tuning series into biomass indices. These time-series of the commercial tuning series were split, in order to better account for changes in catchability due to e.g. technological creep. The age-structured UK(E\&W)-BTS-Q3 survey tuning series was also included. The model was further optimized in terms of parameter configuration for the process and observation models (see table below).

The $\mathrm{F}_{\mathrm{bar}}$ calculates the mean fishing mortality for the set age range and should represent a significant part of the catch. The Fbar in the WGCSE 2019 assessment was set at age 4-8. However, as age 3 represents a large proportion of the catch (Figure 33.3), during the WKFlatNSCS it was decided to expand the $F_{b a r}$ to ages 3-8. The Fbar with ages $3-8$ represents an average $77 \%$ of the catch, with a minimum of $48 \%$ and a maximum of $97 \%$.

The SAM model input and configuration are shown in the table below and in Table 33.10.

| DATA \& SETTINGS |  |
| :---: | :---: |
| tuning indices |  |
| UK(E\&W)-BTS survey (1988-(assessment year-1)) | Age (1-5) |
| BE-CBT_1971-1983 | Biomass |
| BE-CBT_1984-1996 | Biomass |
| BE-CBT3_2006-(assessment year-1) | Biomass |
| UK(E\&W)-CBT_1984-2005 | Biomass |
| UK(E\&W)-CBT_2006-(assessment year-1) | Biomass |
| catch numbers-at-age | Catch numbers for age 1 and 2 set to NA prior 2004 |
| maturity ogive | Age1 = 0; Age2 = 0.67; Age3 = .91; Age4 = .98; Age5 = .99; Age6 = .99; Age6+ = 1 |
| natural mortality | 0.1 for all ages and years |
| prop. M < spawning | 0 for all years |
| prop. F < spawning | 0 for all years |
| Plus group | 10 |
| Fbar | 3-8 |
| MODEL CONFIGURATION |  |
| stock-recruitment | plain random walk on $\operatorname{logN(1)}$ |
| correlation F-at-age | AR(1) |
| F parameters-at-age | $6=0,1,2,3,3,3,4,4,5,5$ |
| q parameters (-at-age) |  |
| UK(E\&W)-BTS survey (1988-(assessment year-1)) | $4=0,1,2,3,3,-1,-1,-1,-1,-1$ |
| BE-CBT_1971-1983 | 1 |
| BE-CBT_1984-1996 | 1 |
| BE-CBT3_2006-(assessment year-1) | 1 |
| UK(E\&W)-CBT_1984-2005 | 1 |
| UK(E\&W)-CBT_2006-(assessment year-1) | 1 |
| $\sigma^{2} \mathrm{~F}$ parameters-at-age | $1=0,0,0,0,0,0,0,0,0,0$ |
| $\sigma^{2} \mathrm{~N}$ parameters-at-age | $2=0,1,1,1,1,1,1,1,1,1$ |
| $\sigma^{\mathbf{2}}$ obs pars (-at-age) |  |
| catch numbers-at-age | $2=0,0,1,1,1,1,1,1,1,1$ |
| UK(E\&W)-BTS survey (1988-(assessment year-1)) | $3=2,3,3,4,4,-1,-1,-1,-1$ |
| BE-CBT_1971-1983 | 1 |
| BE-CBT_1984-1996 | 1 |
| BE-CBT3_2006-(assessment year-1) | 1 |
| UK(E\&W)-CBT_1984-2005 | 1 |
| UK(E\&W)-CBT_2006-(assessment year-1) | 1 |
| p observations at-age |  |
| catch numbers-at-age | "AR(1)" (single $\rho$ for all ages) |
| UK(E\&W)-BTS survey (1988-(assessment year-1)) | "ID" |
| BE-CBT_1971-1983 | - |
| BE-CBT_1984-1996 | - |
| BE-CBT3_2006-(assessment year-1) | - |
| UK(E\&W)-CBT_1984-2005 | - |
| UK(E\&W)-CBT_2006-(assessment year-1) | - |

## This year's assessment

The SAM model fitting diagnostics and survey catchabilities are shown in Table 33.11, the fishing mortalities in Table 31.12, the stock numbers in Table 31.13 and the assessment summary in Table 33.14 and Figure 33.12.

In general, the estimated catches from the SAM model corroborate the observed catches. Mainly at the start of the time-series, some observed catches do not fall within the confidence bounds of the estimated catches. The SAM catch estimate for 2020 is also considerably lower than the ICES catch estimate.

Spawning-stock biomass (SSB) has been above MSY Btrigger since 2009 and shows an increasing trend over the last years, with the 2019, 2020 and 2021 estimates at the same high level. Fishing mortality was below Fmsy in 2017-2019, but increased again and was at Fmsy in 2020 and 2021. Recruitment has been variable without an overall trend. The 2017 recruitment is estimated to be among the highest in the time-series. Recruitment estimates have been above average since 2015.

The one-step ahead residuals for the final SAM assessment are shown in Figure 33.13. There may be some indications of a trend in the UK beam trawl fleet (UK(E\&W)-CBT_2006-2021) with predominantly negative residuals in the last years, in contrast to the positive residuals of the BE_CBT_2006-2021 during the last years.

Retrospective patterns for the final run are shown in Figure 33.14. Retrospective analysis does not indicate major problems; the retrospective patterns are within the confidence bounds. A Mohn's rho analysis was conducted based on the SAM stock assessment results, i.e. the last data year (2021) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The Mohn's rho values for this assessment are low and well within the bounds of $-15 \%$ to $20 \%$ suggested by ICES, i.e. the current assessment indicates sufficient consistency for advice purposes.
The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (AGES 3-8) | RECRUITMENT |
| :--- | :---: | :---: | :---: |
| Mohn's rho value | -0.071 | 0.1 | 0.055 |

## Comparison with previous assessments

The 2021 assessment was revised in October 2021 due to the discovery of an error in the setting of the catch numbers for age 1 and 2 for 1971-2003 in the SAM model. The comparison was done with the updated values.

A comparison of the estimates of this year's assessment with last year's is given in Figure 33.15. Trends in fishing mortality, SSB and recruitment are very similar. The 2018 recruitment was revised upwards by $7.4 \%$ in this year's assessment, whereas the 2019 and 2020 recruitments were revised downwards by $7.7 \%$ and $5 \%$ respectively in this year's assessment. In last year's assessment, F and SSB for 2020 were estimated to be 0.269 and 5811 t respectively; this year's estimates for 2020 are 0.251 and $5945 t$, a downward revision of $6.9 \%$ for $F$ and an upward revision of $2.3 \%$ for SSB.

## State of the stock

Trends in catch, SSB, $\operatorname{Fbar}(3-8)$ and recruitment are presented in Table 33.14 and Figure 33.12.
In the beginning of the time-series, fishing mortality fluctuated around Fmsy ( 0.251 ). During the eighties and nineties fishing mortality increased for this stock to levels well above Fmsy ( 0.51 in 1997). In the following decades, fishing mortality decreased. Fishing mortality has been just below Fmsy in 2017-2019 but slightly increased again and is at Fmš in 2020 and 2021.

Recruitment has fluctuated around 5 million recruits with occasional strong year classes. The 1998 year class is estimated to be among the strongest in the time-series (13 679 thousand fish). Recruitment has been above average ( 5842 thousand fish) since 2015, and the recruitment of 2017 is estimated to be the highest of the time-series (14 305 thousand fish).

SSB has declined almost continuously from the highest value of 6636 t in 1971 to the lowest observed in the time-series in 1997 (2191 t). The exceptional year class of 1998 has increased SSB to above the long-term average ( 3947 t ) in 2022. With the exemption of 2007 and 2008 SSB has been above MSY Btrigger ( 3057 t ) since 2001. SSB increased during the last years, as a result of the decreasing fishing mortality and good recruitment. SSB in 2019-2021 (around 5900 thousand fish) is at its highest level since 1973.

### 33.4 Short-term projections

## Estimating year-class abundance

Higher recruitment was observed from 2015 onwards. However, the 2020 year class ((2021 recruitment) is now estimated at 8071 thousand fish at age 1 (Table 33.13), which is $60 \%$ higher than the median resampled recruitment from the years 1971-2018 (5055 thousand fish), used in last year's forecast.

The age 1 estimates are almost solely coming from the UK(E\&W)-BTS-Q3. In 1999 (24.6), 2008 (10.4) and 2017 (10.5) above average UK(E\&W)-BTS-Q3 abundance indices for age 1 were noted. Those high abundance indices align with the high recruitment in those years. The 2021 UK(E\&W)-BTS-Q3 abundance index for age 1 (9.0) is around the same level of 2000 (9.2).

## Forecast assumptions

The F in $2021(0.25)$ is slightly higher than the mean F over the last three years ( $\mathrm{F}_{\text {average }}$ 2019-2021 $=$ 0.246 ) and using this F to project the stock into 2022, would result in slightly higher landings ( 1357 t ) than the 2022 TAC ( 1337 t ). As recent landings have been close to the TAC (Figure 33.16), the working group decided to use a TAC constraint for the intermediate year (2022). Assuming a TAC constraint for 2022 of 1337 t , implies a fishing mortality in 2022 of 0.246 . This results in an SSB of 6120 t in 2023.

As input for the forecast fishing mortality was calculated as the mean of 2019-2021, scaled to 2021 (0.25). Catch and stock weights-at-age were also averages for the years 2019-2021. Population numbers at the start of 2022 for ages 3 and older, were taken from the SAM output. The long-term median resampled recruitment (1971-2019) as estimated by a stochastic projection (SAM, 5111 thousand fish) was assumed for recruitment in 2022 and subsequent years.

There are no known specific environmental drivers known for this stock.

## MSY forecast

Table 33.15 and Figure 33.17 show the output of the forecast targeting F = Fmsy for 2023-2024 and Figure 33.18 the year classes contributing to the forecast yield and SSB. The assumed long-term median resampled recruitment (1971-2019) accounts for about 4.5\% of the landings in 2023 and about $16.4 \%$ of the 2024 SSB.

Implementing the MSY approach with $\mathrm{F}=\mathrm{FmsY}_{\mathrm{MS}}=0.251$ leads to a total yield of 1338 t in 2023, and an SSB of 5903 t in 2024.

The advice for 2023 is comparable to the advice for 2022 ( $0.075 \%$ advice change).

## Additional options

A management options table is provided in Table 33.15.

### 33.5 Biological reference points

Current biological reference points calculated during the WGCSE 2020 are given in the text table below:

| Framework | Reference point | Value | Technical basis |
| :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 3057 | Tonnes; $\mathrm{B}_{\text {pa }}$ |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.251 | EQsim analysis based on the recruitment period 1971-2018 |
| Precautionary approach | $\mathrm{B}_{\text {lim }}$ | 2184 | Tonnes; $\mathrm{B}_{\text {loss }}$ estimated in 2020, corresponding to SSB in 1997 |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 3057 | Tonnes; $\mathrm{B}_{\mathrm{lim}} \times 1.4$ |
|  | $F_{\text {lim }}$ | 0.543 | EQsim analysis, based on the recruitment period 1971-2018 |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.402 | Fp. $05 ; \mathrm{F}$ that leads to $\mathrm{SSB} \geq \mathrm{Bl}_{\text {lim }}$ with $95 \%$ probability. |
| Management plan | $\underset{\text { ger }}{\text { MAP MSY B }}$ | 3057 | Tonnes; MSY $\mathrm{B}_{\text {trigger }}$ |
|  | MAP $\mathrm{B}_{\mathrm{pa}}$ | 3057 | Tonnes; $\mathrm{B}_{\text {pa }}$ |
|  | MAP $\mathrm{Bl}_{\text {lim }}$ | 2184 | Tonnes; $\mathrm{Bl}_{\text {lim }}$ |
|  | MAP $\mathrm{F}_{\text {MSY }}$ | 0. 251 | $\mathrm{F}_{\text {MSY }}$ |
|  | MAP range <br> $\mathrm{F}_{\text {lower }}$ | 0.136-0.251 | Consistent with ranges provided by ICES (2020), resulting in no more than 5\% reduction in long-term yield compared with MSY |
|  | MAP range $\mathrm{F}_{\text {up- }}$ <br> per | 0.251-0.462 | Consistent with ranges provided by ICES (2020), resulting in no more than 5\% reduction in long-term yield compared with MSY |

[^20]
### 33.6 Management plans

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to demersal stocks including sole in ICES divisions 7.f and 7.g.

### 33.7 Uncertainties and bias in assessment and forecast

## Sampling

The major fleets fishing for 7.f and 7.g sole are sampled (approximately $90 \%$ of the total landings). Sampling is considered to be at a reasonable level.

## Discards

Discard estimates used to be low, but are increasing. Discards are included in the assessment since the WGCSE 2020.

## Misreporting

Area misreporting is known to have been considerable over the period 2002-2005. This was due to a combination of the good 1998 year class still being an important part of the catch composition and more restrictive TACs. The area misreporting has been corrected for the years 2002-2006 (WGSSDS 2007). At the WKCELT 2014, analysis revealed that there was additional misreporting taking place in 2002-2003 and 2004, which was not accounted for in the first correction done at WGSSDS in 2007. Since 2007, the area misreporting that could be estimated was negligible. During the WKFlatNSCS (ICES, 2020) a further correction for 2004-2007 landings data was done.

## Surveys

The UK(E\&W)-BTS-Q3 survey, which is solely responsible for the recruiting estimates, has been able to track year-class strength at-ages greater than 0 rather well in the past. However, the estimates of strong year classes have sometimes been revised downward in previous assessments and may cause bias in the forecast.

### 33.8 Recommendations for next Benchmark

Sole in 7.f and 7.g have been benchmarked in February 2020. The remaining issues are listed below.

| Problem / Aim | Work needed / Work <br> needed / possible direc- <br> tion of solution | Data needed to be <br> able to do this: are <br> these available / <br> where should these <br> come from? |  |
| :--- | :--- | :--- | :--- |
| Natural mortality |  | *estimates of natural mortal- | *estimates of natural |
| Alternate rates of natural mortality. Natural mortality is assumed | ity |  |  |
| constant over ages and years at 0.1. When new information is availa- |  |  |  |

### 33.9 Management considerations

The stock-recruitment relationship is not well-defined, there is no real evidence of reduced recruitment at low levels of SSB for this (Figure 33.19). Following the recent strong year classes, SSB is now at its highest level since 1973.

The Celtic Sea is an area without days-at-sea limitations for demersal fisheries. In this context and given that many demersal vessels are very mobile, changes in effort measures in areas other than the Celtic Sea, can influence the effort regime in the Celtic Sea (cfr. increased effort in Celtic Sea for Belgian beamers during 2004-2005 when days-at-sea limitations were in place for the Eastern English Channel).

### 33.10 Ecosystem considerations

Sole and plaice are predominantly caught by beam trawl fisheries. Beam trawling is known to have an impact on the benthic communities, although less so on soft substrates and in areas which have been historically exploited by this fishing method. Benthic drop-out panels have been shown to release around $75 \%$ of benthic invertebrates from the catches. Information from the UK industry (Trebilcock and Rozarieux, 2009) suggests that uptake in 2008 was minimal.

A complete ecosystem overview can be found in the stock annex section A.3.

### 33.11 References

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Table 33.1. Sol.27.7fg - Official Nominal landings and landings and discard data used by the Working Group (t).

| Year | Belgium |  | Denmark | France | Ireland | UK(E.\&W, NI) | UK(Scotland) | Spain | Other | Total-Official | Unallocated | Used by WG | TAC | Discards** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 1039 | * | 2 | 146 | 188 | 611 | - | - | 3 | 1989 | -389 | 1600 |  | 80 |
| 1987 | 701 | * | - | 117 | 9 | 437 | - | - | - | 1264 | -42 | 1222 | 1600 | 56 |
| 1988 | 705 | * | - | 110 | 72 | 317 | - | - | - | 1204 | -58 | 1146 | 1100 | 61 |
| 1989 | 684 | * | - | 87 | 18 | 203 | - | - | - | 992 | 0 | 992 | 1000 | 70 |
| 1990 | 716 | * | - | 130 | 40 | 353 | 0 | - | - | 1239 | -50 | 1189 | 1200 | 57 |
| 1991 | 982 | * | - | 80 | 32 | 402 | 0 | - | - | 1496 | -389 | 1107 | 1200 | 126 |
| 1992 | 543 | * | - | 141 | 45 | 325 | 6 | - | - | 1060 | -79 | 981 | 1200 | 77 |
| 1993 | 575 | * | - | 108 | 51 | 285 | 11 | - | - | 1030 | -102 | 928 | 1100 | 56 |
| 1994 | 619 | * | - | 90 | 37 | 264 | 8 | - | - | 1018 | -9 | 1009 | 1100 | 52 |
| 1995 | 763 | * | - | 88 | 20 | 294 | - | - | - | 1165 | -8 | 1157 | 1100 | 50 |
| 1996 | 695 | * | - | 102 | 19 | 265 | 0 | - | - | 1081 | -86 | 995 | 1000 | 47 |
| 1997 | 660 | * | - | 99 | 28 | 251 | 0 | - | - | 1038 | -111 | 927 | 900 | 46 |
| 1998 | 675 | * | - | 98 | 42 | 198 | - | - | - | 1013 | -138 | 875 | 850 | 43 |
| 1999 | 604 |  | - | 61 | 51 | 231 | 0 | - | - | 947 | 65 | 1012 | 960 | 89 |
| 2000 | 694 |  | - | 74 | 29 | 243 | - | - | - | 1040 | 51 | 1091 | 1160 | 158 |
| 2001 | 720 |  | - | 77 | 35 | 288 | - | - | - | 1120 | 48 | 1168 | 1020 | 101 |
| 2002 | 703 |  | - | 65 | 32 | 318 | + | - | - | 1118 | 227 | 1345 | 1070 | 58 |
| 2003 | 715 |  | - | 124 | 26 | 342 | + | - | - | 1207 | 340 | 1547 | 1240 | 54 |


| Year | Belgium | Denmark | France | Ireland | UK(E.\&W, NI) | UK(Scotland) | Spain | Other | Total-Official | Unallocated | Used by WG | TAC | Discards** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 735 | - | 79 | 33 | 283 | - | - | - | 1130 | 261 | 1391 | 1050 | 140 |
| 2005 | 645 | - | 101 | 34 | 217 | - | - | - | 997 | 266 | 1263 | 1000 | 23 |
| 2006 | 576 | - | 75 | 38 | 232 | - | - | - | 921 | 137 | 1058 | 950 | 41 |
| 2007 | 582 | - | 85 | 32 | 245 | - | - | - | 943 | 109 | 1052 | 890 | 36 |
| 2008 | 466 | - | 68 | 28 | 218 | - | - | - | 781 | 9 | 790 | 964 | 8 |
| 2009 | 513 | - | 73 | 26 | 195 | - | - | - | 806 | -34 | 772 | 993 | 30 |
| 2010 | 620 | - | 44 | 27 | 180 | - | - | - | 871 | -4 | 867 | 993 | 56 |
| 2011 | 775 | - | 54 | 30 | 168 | - | - | - | 1027 | 0 | 1027 | 1241 | 28 |
| 2012 | 843 | - | 48 | 33 | 175 | - | - | - | 1099 | 2 | 1101 | 1060 | 32 |
| 2013 | 789 | - | 49 | 42 | 205 | - | - | - | 1085 | 8 | 1093 | 1100 | 26 |
| 2014 | 703 | - | 58 | 28 | 252 | - | - | - | 1042 | -1 | 1041 | 1001 | 27 |
| 2015 | 674 | - | 24 | 27 | 105 | - | - | - | 830 | 1 | 831 | 851 | 17 |
| 2016 | 563 | - | 72 | 21 | 174 | - | - | - | 830 | 2 | 832 | 779 | 31 |
| 2017 | 551 | - | 49 | 28 | 149 | - | - | - | 777 | 1 | 778 | 845 | 65 |
| 2018 | 607 | - | 44 | 27 | 171 | - | - | - | 850 | 0 | 850 | 920 | 141 |
| 2019 | 800 | - | 42 | 33 | 193 | - | <1 | - | 1068 | 0 | 1068 | 1009 | 145 |
| 2020 ^ | 1121 | - | 44 | 51 | 291 | - | <1 | - | 1507 | 17 | 1524 | 1652 | 106 |
| 2021 ^ | 859 | - | 61 | 48 | 395 | - | 2 | - | 1365 | -29 | 1336 | 1337 | 62 |

${ }^{\wedge}$ Landings are preliminary.

* including 7.g-k.
** Discards estimated by ICES.

Table 33.2. Sol.27.7fg - Catch numbers-at-age (in thousands).

| year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0 | 610 | 303 | 1377 | 638 | 439 | 541 | 770 | 379 | 1231 | 6288 |
| 1972 | 0 | 855 | 1014 | 322 | 684 | 334 | 214 | 234 | 317 | 739 | 4713 |
| 1973 | 0 | 575 | 2116 | 768 | 311 | 357 | 120 | 111 | 117 | 649 | 5124 |
| 1974 | 0 | 245 | 492 | 886 | 420 | 212 | 241 | 98 | 110 | 547 | 3251 |
| 1975 | 0 | 188 | 323 | 345 | 652 | 308 | 111 | 103 | 68 | 375 | 2473 |
| 1976 | 0 | 493 | 937 | 575 | 624 | 567 | 263 | 132 | 199 | 469 | 4259 |
| 1977 | 0 | 496 | 492 | 358 | 277 | 248 | 407 | 121 | 28 | 368 | 2795 |
| 1978 | 0 | 502 | 833 | 348 | 157 | 161 | 100 | 200 | 72 | 175 | 2548 |
| 1979 | 0 | 519 | 630 | 767 | 212 | 156 | 198 | 125 | 154 | 170 | 2931 |
| 1980 | 0 | 1038 | 1092 | 899 | 596 | 183 | 62 | 97 | 101 | 355 | 4423 |
| 1981 | 0 | 951 | 759 | 813 | 407 | 382 | 151 | 121 | 95 | 383 | 4062 |
| 1982 | 0 | 540 | 934 | 317 | 477 | 284 | 208 | 93 | 112 | 328 | 3293 |
| 1983 | 0 | 1023 | 1212 | 748 | 290 | 354 | 227 | 194 | 52 | 322 | 4422 |
| 1984 | 0 | 1062 | 951 | 622 | 553 | 187 | 279 | 107 | 47 | 276 | 4084 |
| 1985 | 0 | 310 | 1656 | 786 | 577 | 300 | 101 | 141 | 74 | 241 | 4186 |
| 1986 | 0 | 781 | 1457 | 1204 | 537 | 363 | 194 | 88 | 104 | 330 | 5058 |
| 1987 | 0 | 503 | 1076 | 818 | 589 | 277 | 206 | 101 | 61 | 180 | 3811 |
| 1988 | 0 | 831 | 522 | 902 | 450 | 393 | 128 | 79 | 68 | 270 | 3643 |
| 1989 | 0 | 757 | 1308 | 617 | 634 | 240 | 189 | 83 | 24 | 102 | 3954 |
| 1990 | 0 | 438 | 1117 | 1207 | 407 | 459 | 139 | 116 | 50 | 130 | 4063 |
| 1991 | 0 | 2304 | 776 | 676 | 507 | 153 | 157 | 56 | 46 | 163 | 4838 |
| 1992 | 0 | 684 | 1911 | 661 | 418 | 257 | 61 | 60 | 28 | 89 | 4169 |
| 1993 | 0 | 559 | 970 | 1133 | 339 | 189 | 162 | 64 | 84 | 99 | 3599 |
| 1994 | 0 | 466 | 888 | 759 | 882 | 287 | 150 | 66 | 42 | 146 | 3686 |
| 1995 | 0 | 204 | 1299 | 1127 | 429 | 490 | 134 | 113 | 66 | 109 | 3971 |
| 1996 | 0 | 280 | 1163 | 928 | 433 | 232 | 193 | 58 | 43 | 106 | 3436 |
| 1997 | 0 | 387 | 1000 | 615 | 408 | 256 | 128 | 127 | 45 | 106 | 3072 |
| 1998 | 0 | 311 | 1048 | 743 | 303 | 173 | 109 | 51 | 52 | 87 | 2877 |
| 1999 | 0 | 961 | 1931 | 856 | 288 | 145 | 81 | 31 | 23 | 44 | 4360 |
| 2000 | 0 | 2720 | 1664 | 701 | 246 | 61 | 56 | 43 | 19 | 51 | 5561 |
| 2001 | 0 | 1111 | 2155 | 883 | 445 | 245 | 65 | 39 | 26 | 81 | 5050 |
| 2002 | 0 | 46 | 1647 | 2261 | 674 | 253 | 96 | 55 | 36 | 51 | 5119 |
| 2003 | 0 | 209 | 871 | 1294 | 2111 | 453 | 250 | 90 | 29 | 84 | 5391 |
| 2004 | 5 | 393 | 1846 | 941 | 1086 | 742 | 132 | 100 | 54 | 100 | 5399 |
| 2005 | 12 | 418 | 1096 | 1028 | 592 | 499 | 336 | 72 | 55 | 89 | 4197 |
| $\underline{2006}$ | 8 | 485 | 1151 | 844 | 706 | 250 | 229 | 169 | 60 | 106 | 4008 |
| 2007 | 19 | 697 | 979 | 721 | 435 | 382 | 149 | 142 | 155 | 93 | 3772 |
| 2008 | 10 | 180 | 515 | 499 | 387 | 212 | 209 | 85 | 109 | 150 | 2356 |
| $\underline{2009}$ | 30 | 549 | 511 | 588 | 435 | 259 | 164 | 121 | 51 | 203 | 2911 |
| 2010 | 26 | 506 | 1510 | 657 | 380 | 257 | 140 | 103 | 80 | 119 | 3778 |
| 2011 | 18 | 170 | 1103 | 1389 | 394 | 308 | 187 | 118 | 56 | 168 | 3911 |
| $\underline{2012}$ | 29 | 361 | 318 | 1039 | 1339 | 370 | 222 | 130 | 84 | 219 | 4111 |
| 2013 | 13 | 545 | 998 | 523 | 826 | 652 | 222 | 104 | 61 | 160 | 4104 |
| 2014 | 14 | 173 | 684 | 735 | 308 | 388 | 381 | 122 | 99 | 243 | 3147 |
| $\underline{2015}$ | 0 | 193 | 837 | 924 | 433 | 145 | 201 | 114 | 69 | 113 | 3029 |
| $\underline{2016}$ | 66 | 727 | 458 | 635 | 663 | 303 | 111 | 132 | 94 | 70 | 3259 |
| $\underline{2017}$ | 47 | 432 | 1157 | 493 | 421 | 353 | 147 | 55 | 59 | 101 | 3265 |
| $\underline{2018}$ | 4 | 989 | 840 | 1105 | 275 | 293 | 186 | 95 | 56 | 122 | 3965 |
| $\underline{2019}$ | 53 | 373 | 2240 | 729 | 874 | 306 | 162 | 115 | 105 | 118 | 5075 |
| $\underline{2020}$ | 0 | 242 | 777 | 2775 | 1161 | 789 | 304 | 160 | 106 | 189 | 6503 |
| $\underline{2021}$ | 0 | 128 | 688 | 1077 | 1944 | 942 | 361 | 126 | 63 | 171 | 5500 |

Table 33.3. Sol.27.7fg - Catch weights-at-age (kg).

| year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0.039 | 0.11 | 0.168 | 0.224 | 0.273 | 0.316 | 0.353 | 0.384 | 0.408 | 0.441 |
| 1972 | 0.106 | 0.136 | 0.185 | 0.227 | 0.265 | 0.303 | 0.34 | 0.377 | 0.413 | 0.539 |
| 1973 | 0.081 | 0.134 | 0.2 | 0.259 | 0.311 | 0.361 | 0.408 | 0.452 | 0.493 | 0.602 |
| 1974 | 0.063 | 0.13 | 0.202 | 0.27 | 0.329 | 0.385 | 0.436 | 0.483 | 0.524 | 0.624 |
| 1975 | 0.046 | 0.127 | 0.208 | 0.286 | 0.355 | 0.416 | 0.473 | 0.523 | 0.565 | 0.671 |
| 1976 | 0.114 | 0.149 | 0.214 | 0.268 | 0.316 | 0.363 | 0.409 | 0.453 | 0.496 | 0.665 |
| 1977 | 0.098 | 0.15 | 0.229 | 0.297 | 0.355 | 0.408 | 0.46 | 0.506 | 0.548 | 0.668 |
| 1978 | 0.068 | 0.141 | 0.228 | 0.308 | 0.377 | 0.44 | 0.498 | 0.55 | 0.596 | 0.72 |
| 1979 | 0.023 | 0.127 | 0.226 | 0.32 | 0.4 | 0.47 | 0.531 | 0.58 | 0.621 | 0.664 |
| 1980 | 0.048 | 0.134 | 0.228 | 0.315 | 0.391 | 0.459 | 0.523 | 0.578 | 0.625 | 0.72 |
| 1981 | 0.078 | 0.141 | 0.22 | 0.292 | 0.355 | 0.413 | 0.469 | 0.519 | 0.564 | 0.665 |
| 1982 | 0.061 | 0.142 | 0.236 | 0.323 | 0.396 | 0.461 | 0.521 | 0.571 | 0.616 | 0.704 |
| 1983 | 0.085 | 0.153 | 0.247 | 0.329 | 0.397 | 0.458 | 0.513 | 0.56 | 0.602 | 0.679 |
| 1984 | 0.019 | 0.126 | 0.229 | 0.329 | 0.414 | 0.492 | 0.561 | 0.621 | 0.673 | 0.771 |
| 1985 | 0.089 | 0.151 | 0.239 | 0.316 | 0.382 | 0.443 | 0.499 | 0.551 | 0.596 | 0.703 |
| 1986 | 0.046 | 0.134 | 0.23 | 0.32 | 0.399 | 0.47 | 0.536 | 0.593 | 0.643 | 0.748 |
| 1987 | 0.048 | 0.136 | 0.23 | 0.319 | 0.395 | 0.465 | 0.528 | 0.583 | 0.632 | 0.74 |
| 1988 | 0.074 | 0.143 | 0.229 | 0.308 | 0.377 | 0.441 | 0.502 | 0.556 | 0.606 | 0.738 |
| 1989 | 0.013 | 0.112 | 0.196 | 0.28 | 0.355 | 0.423 | 0.487 | 0.542 | 0.592 | 0.691 |
| 1990 | 0.049 | 0.128 | 0.21 | 0.291 | 0.362 | 0.429 | 0.494 | 0.552 | 0.609 | 0.747 |
| 1991 | 0.054 | 0.138 | 0.232 | 0.319 | 0.392 | 0.458 | 0.516 | 0.564 | 0.608 | 0.674 |
| 1992 | 0.073 | 0.136 | 0.212 | 0.281 | 0.342 | 0.397 | 0.451 | 0.499 | 0.543 | 0.64 |
| 1993 | 0.057 | 0.128 | 0.204 | 0.275 | 0.338 | 0.396 | 0.45 | 0.5 | 0.544 | 0.645 |
| 1994 | 0.081 | 0.139 | 0.212 | 0.276 | 0.331 | 0.38 | 0.425 | 0.465 | 0.5 | 0.563 |
| 1995 | 0.068 | 0.136 | 0.216 | 0.288 | 0.351 | 0.408 | 0.462 | 0.51 | 0.552 | 0.643 |
| 1996 | 0.027 | 0.122 | 0.21 | 0.296 | 0.371 | 0.438 | 0.5 | 0.551 | 0.598 | 0.677 |
| 1997 | 0.074 | 0.142 | 0.228 | 0.306 | 0.375 | 0.439 | 0.5 | 0.554 | 0.605 | 0.707 |
| 1998 | 0.079 | 0.147 | 0.237 | 0.319 | 0.392 | 0.461 | 0.527 | 0.589 | 0.647 | 0.781 |
| 1999 | 0.015 | 0.121 | 0.217 | 0.314 | 0.399 | 0.476 | 0.548 | 0.613 | 0.67 | 0.765 |
| 2000 | 0.078 | 0.148 | 0.24 | 0.321 | 0.389 | 0.45 | 0.506 | 0.553 | 0.594 | 0.665 |
| 2001 | 0.066 | 0.137 | 0.22 | 0.296 | 0.362 | 0.424 | 0.482 | 0.533 | 0.579 | 0.677 |
| 2002 | 0.054 | 0.126 | 0.2 | 0.271 | 0.336 | 0.398 | 0.457 | 0.512 | 0.564 | 0.704 |
| 2003 | 0.123 | 0.151 | 0.214 | 0.266 | 0.313 | 0.361 | 0.408 | 0.454 | 0.501 | 0.639 |
| 2004 | 0.099 | 0.152 | 0.194 | 0.274 | 0.347 | 0.371 | 0.459 | 0.522 | 0.524 | 0.631 |
| 2005 | 0.109 | 0.155 | 0.203 | 0.267 | 0.346 | 0.439 | 0.473 | 0.595 | 0.624 | 0.707 |
| 2006 | 0.142 | 0.144 | 0.186 | 0.272 | 0.33 | 0.401 | 0.412 | 0.411 | 0.465 | 0.574 |
| 2007 | 0.161 | 0.157 | 0.221 | 0.284 | 0.335 | 0.372 | 0.414 | 0.488 | 0.511 | 0.589 |
| 2008 | 0.149 | 0.163 | 0.21 | 0.281 | 0.361 | 0.359 | 0.449 | 0.62 | 0.625 | 0.59 |
| 2009 | 0.105 | 0.157 | 0.188 | 0.242 | 0.294 | 0.348 | 0.378 | 0.476 | 0.485 | 0.546 |
| $\underline{2010}$ | 0.11 | 0.144 | 0.179 | 0.234 | 0.312 | 0.358 | 0.387 | 0.414 | 0.524 | 0.616 |
| 2011 | 0.123 | 0.155 | 0.185 | 0.233 | 0.277 | 0.361 | 0.431 | 0.465 | 0.483 | 0.688 |
| 2012 | 0.102 | 0.142 | 0.206 | 0.243 | 0.271 | 0.312 | 0.35 | 0.38 | 0.417 | 0.543 |
| 2013 | 0.092 | 0.158 | 0.195 | 0.249 | 0.29 | 0.329 | 0.361 | 0.463 | 0.492 | 0.587 |
| 2014 | 0.138 | 0.15 | 0.199 | 0.268 | 0.347 | 0.394 | 0.427 | 0.496 | 0.523 | 0.702 |
| 2015 | 0.135 | 0.132 | 0.193 | 0.248 | 0.311 | 0.367 | 0.438 | 0.502 | 0.463 | 0.589 |
| 2016 | 0.109 | 0.165 | 0.203 | 0.251 | 0.285 | 0.342 | 0.416 | 0.429 | 0.47 | 0.621 |
| 2017 | 0.093 | 0.151 | 0.2 | 0.255 | 0.315 | 0.33 | 0.382 | 0.444 | 0.513 | 0.549 |
| 2018 | 0.082 | 0.175 | 0.187 | 0.251 | 0.294 | 0.327 | 0.385 | 0.436 | 0.425 | 0.575 |
| 2019 | 0.099 | 0.12 | 0.17 | 0.24 | 0.307 | 0.327 | 0.419 | 0.451 | 0.445 | 0.614 |
| 2020 | 0.155 | 0.114 | 0.176 | 0.205 | 0.257 | 0.331 | 0.363 | 0.415 | 0.504 | 0.556 |
| 2021 | 0.12 | 0.127 | 0.17 | 0.212 | 0.252 | 0.26 | 0.35 | 0.412 | 0.493 | 0.537 |

Table 33.4. - Sol.27.7fg - Annual landings length distributions by fleet.

|  | UK (England \& Wales) | Belgium | Ireland |
| :---: | :---: | :---: | :---: |
| Length (cm) | Beam trawl |  |  |
| 17 |  |  |  |
| 18 |  |  |  |
| 19 |  |  |  |
| 20 |  |  |  |
| 21 |  | 192 |  |
| 22 | 9 | 2619 |  |
| 23 | 3849 | 18609 |  |
| 24 | 10514 | 111613 | 16 |
| 25 | 29791 | 365739 | 844 |
| 26 | 50277 | 460218 | 1297 |
| 27 | 69733 | 449261 | 1785 |
| 28 | 85621 | 415013 | 3713 |
| 29 | 114396 | 326791 | 4268 |
| 30 | 111213 | 287587 | 4865 |
| 31 | 126187 | 195077 | 3576 |
| 32 | 93874 | 150824 | 5044 |
| 33 | 81263 | 117136 | 4258 |
| 34 | 59104 | 92599 | 3066 |
| 35 | 82402 | 71869 | 2801 |
| 36 | 51647 | 54728 | 2585 |
| 37 | 36187 | 35589 | 1718 |
| 38 | 25149 | 27470 | 1210 |
| 39 | 19331 | 20718 | 1192 |
| 40 | 13171 | 17462 | 1035 |
| 41 | 9421 | 6205 | 1048 |
| 42 | 5379 | 5207 | 273 |
| 43 | 7088 | 2847 | 228 |
| 44 | 2612 | 3102 | 434 |
| 45 | 1911 | 2544 | 233 |
| 46 | 539 | 1781 | 254 |
| 47 | 1442 | 356 |  |
| 48 | 423 | 652 |  |
| 49 | 137 | 241 |  |
| 50 | 91 | 195 |  |
| 51 | 11 |  |  |
| 52 |  |  |  |
| 53 | 11 |  |  |
| 54 |  |  |  |
| 55 |  |  |  |
| 56 |  |  |  |
| 57 |  |  |  |
| 58 |  |  |  |
| 59 |  |  |  |
| 60 |  |  |  |
| Total | 1092783 | 3244244 | 45743 |

Table 33.5. Sol.27.7fg - Stock weights-at-age (kg).

| year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0.041 | 0.085 | 0.145 | 0.205 | 0.26 | 0.304 | 0.341 | 0.37 | 0.39 | 0.416 |
| 1972 | 0.041 | 0.073 | 0.143 | 0.196 | 0.244 | 0.288 | 0.328 | 0.365 | 0.398 | 0.511 |
| 1973 | 0.041 | 0.119 | 0.165 | 0.219 | 0.266 | 0.31 | 0.352 | 0.392 | 0.431 | 0.56 |
| 1974 | 0.041 | 0.103 | 0.165 | 0.232 | 0.292 | 0.346 | 0.397 | 0.444 | 0.487 | 0.599 |
| 1975 | 0.041 | 0.089 | 0.165 | 0.241 | 0.31 | 0.37 | 0.426 | 0.477 | 0.523 | 0.649 |
| 1976 | 0.041 | 0.083 | 0.165 | 0.237 | 0.301 | 0.359 | 0.413 | 0.463 | 0.509 | 0.669 |
| 1977 | 0.041 | 0.131 | 0.185 | 0.252 | 0.308 | 0.359 | 0.409 | 0.455 | 0.498 | 0.643 |
| 1978 | 0.041 | 0.118 | 0.185 | 0.266 | 0.335 | 0.395 | 0.451 | 0.503 | 0.549 | 0.695 |
| 1979 | 0.041 | 0.093 | 0.179 | 0.27 | 0.351 | 0.421 | 0.483 | 0.538 | 0.585 | 0.654 |
| 1980 | 0.041 | 0.056 | 0.17 | 0.267 | 0.354 | 0.429 | 0.496 | 0.554 | 0.602 | 0.7 |
| 1981 | 0.041 | 0.082 | 0.172 | 0.258 | 0.334 | 0.402 | 0.464 | 0.521 | 0.571 | 0.674 |
| 1982 | 0.041 | 0.105 | 0.182 | 0.267 | 0.34 | 0.404 | 0.464 | 0.517 | 0.565 | 0.682 |
| 1983 | 0.041 | 0.097 | 0.187 | 0.279 | 0.358 | 0.426 | 0.486 | 0.54 | 0.586 | 0.676 |
| 1984 | 0.041 | 0.104 | 0.187 | 0.285 | 0.369 | 0.442 | 0.507 | 0.564 | 0.614 | 0.733 |
| 1985 | 0.041 | 0.054 | 0.174 | 0.269 | 0.354 | 0.428 | 0.496 | 0.556 | 0.608 | 0.716 |
| 1986 | 0.041 | 0.109 | 0.186 | 0.276 | 0.355 | 0.424 | 0.487 | 0.544 | 0.596 | 0.715 |
| 1987 | 0.041 | 0.079 | 0.176 | 0.271 | 0.355 | 0.43 | 0.498 | 0.559 | 0.612 | 0.737 |
| 1988 | 0.041 | 0.083 | 0.176 | 0.266 | 0.347 | 0.417 | 0.483 | 0.542 | 0.594 | 0.729 |
| 1989 | 0.041 | 0.091 | 0.167 | 0.253 | 0.331 | 0.4 | 0.463 | 0.521 | 0.574 | 0.686 |
| 1990 | 0.041 | 0.041 | 0.154 | 0.239 | 0.319 | 0.39 | 0.457 | 0.519 | 0.575 | 0.721 |
| 1991 | 0.041 | 0.082 | 0.173 | 0.259 | 0.338 | 0.407 | 0.47 | 0.528 | 0.58 | 0.673 |
| 1992 | 0.041 | 0.086 | 0.171 | 0.256 | 0.33 | 0.395 | 0.454 | 0.507 | 0.554 | 0.648 |
| 1993 | 0.041 | 0.097 | 0.167 | 0.242 | 0.308 | 0.368 | 0.423 | 0.475 | 0.521 | 0.631 |
| 1994 | 0.041 | 0.089 | 0.165 | 0.237 | 0.302 | 0.358 | 0.41 | 0.458 | 0.5 | 0.578 |
| 1995 | 0.041 | 0.105 | 0.173 | 0.247 | 0.311 | 0.368 | 0.419 | 0.466 | 0.507 | 0.602 |
| 1996 | 0.041 | 0.091 | 0.169 | 0.253 | 0.327 | 0.392 | 0.452 | 0.504 | 0.552 | 0.647 |
| 1997 | 0.041 | 0.062 | 0.167 | 0.254 | 0.333 | 0.404 | 0.468 | 0.526 | 0.577 | 0.681 |
| 1998 | 0.041 | 0.104 | 0.184 | 0.27 | 0.346 | 0.416 | 0.481 | 0.542 | 0.599 | 0.732 |
| 1999 | 0.041 | 0.098 | 0.179 | 0.273 | 0.357 | 0.432 | 0.502 | 0.569 | 0.628 | 0.739 |
| 2000 | 0.041 | 0.047 | 0.17 | 0.264 | 0.35 | 0.423 | 0.491 | 0.55 | 0.603 | 0.695 |
| 2001 | 0.041 | 0.103 | 0.181 | 0.267 | 0.341 | 0.406 | 0.466 | 0.519 | 0.566 | 0.664 |
| 2002 | 0.041 | 0.091 | 0.165 | 0.244 | 0.315 | 0.38 | 0.44 | 0.497 | 0.548 | 0.688 |
| 2003 | 0.041 | 0.09 | 0.164 | 0.231 | 0.292 | 0.348 | 0.403 | 0.456 | 0.507 | 0.645 |
| 2004 | 0.041 | 0.137 | 0.172 | 0.242 | 0.304 | 0.341 | 0.407 | 0.462 | 0.488 | 0.623 |
| 2005 | 0.041 | 0.124 | 0.175 | 0.228 | 0.308 | 0.391 | 0.419 | 0.523 | 0.571 | 0.668 |
| 2006 | 0.041 | 0.125 | 0.17 | 0.235 | 0.297 | 0.373 | 0.425 | 0.441 | 0.526 | 0.605 |
| 2007 | 0.041 | 0.149 | 0.178 | 0.23 | 0.302 | 0.35 | 0.408 | 0.448 | 0.458 | 0.573 |
| 2008 | 0.041 | 0.162 | 0.181 | 0.249 | 0.32 | 0.347 | 0.409 | 0.507 | 0.552 | 0.569 |
| 2009 | 0.041 | 0.153 | 0.175 | 0.226 | 0.287 | 0.355 | 0.368 | 0.462 | 0.548 | 0.572 |
| 2010 | 0.041 | 0.123 | 0.168 | 0.21 | 0.275 | 0.325 | 0.367 | 0.396 | 0.499 | 0.567 |
| 2011 | 0.041 | 0.13 | 0.163 | 0.204 | 0.254 | 0.336 | 0.393 | 0.424 | 0.447 | 0.651 |
| 2012 | 0.041 | 0.132 | 0.179 | 0.212 | 0.251 | 0.294 | 0.356 | 0.405 | 0.44 | 0.574 |
| 2013 | 0.041 | 0.127 | 0.167 | 0.227 | 0.266 | 0.299 | 0.336 | 0.403 | 0.432 | 0.554 |
| 2014 | 0.041 | 0.118 | 0.177 | 0.229 | 0.294 | 0.338 | 0.375 | 0.423 | 0.492 | 0.668 |
| 2015 | 0.041 | 0.135 | 0.171 | 0.222 | 0.288 | 0.357 | 0.415 | 0.463 | 0.479 | 0.596 |
| 2016 | 0.041 | 0.149 | 0.164 | 0.221 | 0.266 | 0.326 | 0.391 | 0.434 | 0.486 | 0.599 |
| 2017 | 0.041 | 0.128 | 0.182 | 0.227 | 0.281 | 0.306 | 0.361 | 0.43 | 0.469 | 0.544 |
| 2018 | 0.041 | 0.128 | 0.168 | 0.224 | 0.274 | 0.321 | 0.356 | 0.408 | 0.434 | 0.558 |
| 2019 | 0.041 | 0.099 | 0.172 | 0.212 | 0.277 | 0.31 | 0.37 | 0.417 | 0.44 | 0.571 |
| 2020 | 0.041 | 0.106 | 0.145 | 0.187 | 0.249 | 0.319 | 0.344 | 0.417 | 0.477 | 0.555 |
| 2021 | 0.041 | 0.14 | 0.139 | 0.193 | 0.227 | 0.259 | 0.34 | 0.387 | 0.452 | 0.544 |

Table 33.6. Sol.27.7fg - Indices of abundance (No/100km) for UK(E\&W)-BTS-Q3.

| year/age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 30 | 81 | 326 | 49 | 19 | 5 | 0 | 0 | 0 | 0 |
| 1989 | 144 | 222 | 331 | 176 | 20 | 15 | 7 | 4 | 2 | 2 |
| 1990 | 30 | 385 | 313 | 50 | 16 | 4 | 7 | 3 | 0 | 0 |
| 1991 | 32 | 241 | 517 | 67 | 17 | 15 | 4 | 0 | 2 | 2 |
| 1992 | 4 | 394 | 260 | 139 | 30 | 18 | 10 | 1 | 2 | 1 |
| 1993 | 3 | 169 | 320 | 43 | 19 | 1 | 2 | 2 | 1 | 1 |
| 1994 | 1 | 333 | 387 | 99 | 14 | 7 | 7 | 0 | 0 | 2 |
| 1995 | 27 | 124 | 222 | 52 | 11 | 6 | 12 | 1 | 1 | 1 |
| 1996 | 3 | 150 | 212 | 54 | 23 | 6 | 3 | 3 | 1 | 3 |
| 1997 | 32 | 433 | 180 | 18 | 11 | 12 | 4 | 3 | 5 | 0 |
| 1998 | 91 | 770 | 411 | 50 | 10 | 8 | 4 | 2 | 1 | 4 |
| 1999 | 24 | 2464 | 250 | 32 | 13 | 6 | 3 | 4 | 1 | 0 |
| $\underline{2000}$ | 13 | 916 | 1356 | 31 | 22 | 5 | 0 | 2 | 1 | 1 |
| 2001 | 22 | 379 | 600 | 259 | 19 | 8 | 5 | 2 | 0 | 2 |
| 2002 | 8 | 663 | 239 | 127 | 102 | 12 | 6 | 2 | 3 | 0 |
| 2003 | 12 | 392 | 530 | 46 | 25 | 47 | 8 | 3 | 3 | 0 |
| 2004 | 56 | 749 | 378 | 86 | 13 | 19 | 37 | 3 | 3 | 0 |
| 2005 | 37 | 343 | 225 | 32 | 13 | 6 | 4 | 14 | 1 | 2 |
| 2006 | 11 | 273 | 201 | 40 | 13 | 7 | 0 | 2 | 10 | 0 |
| 2007 | 91 | 358 | 108 | 43 | 13 | 7 | 6 | 3 | 3 | 11 |
| 2008 | 5 | 1039 | 105 | 13 | 15 | 6 | 8 | 3 | 3 | 4 |
| 2009 | 1 | 509 | 318 | 24 | 7 | 8 | 3 | 3 | 3 | 2 |
| 2010 | 18 | 85 | 471 | 121 | 17 | 2 | 4 | 8 | 3 | 2 |
| 2011 | 18 | 502 | 52 | 138 | 69 | 7 | 2 | 6 | 3 | 0 |
| 2012 | 13 | 542 | 231 | 8 | 53 | 24 | 1 | 1 | 1 | 3 |
| 2013 | 9 | 279 | 518 | 43 | 13 | 24 | 15 | 1 | 5 | 1 |
| 2014 | 34 | 244 | 257 | 76 | 13 | 5 | 23 | 8 | 1 | 1 |
| $\underline{2015}$ | 28 | 746 | 48 | 44 | 31 | 7 | 3 | 13 | 6 | 0 |
| 2016 | 26 | 573 | 359 | 12 | 27 | 13 | 7 | 3 | 5 | 8 |
| 2017 | 6 | 1046 | 174 | 67 | 13 | 16 | 17 | 4 | 3 | 11 |
| 2018 | 27 | 434 | 906 | 279 | 45 | 17 | 9 | 15 | 11 | 4 |
| 2019 | 2 | 708 | 325 | 164 | 23 | 29 | 3 | 6 | 7 | 4 |
| 2020 | 3 | 331 | 238 | 74 | 67 | 24 | 17 | 2 | 6 | 7 |
| 2021 | 57 | 896 | 154 | 45 | 37 | 33 | 12 | 12 | 0 | 3 |
| Mean | 26.98 | 522.73 | 338.88 | 76.51 | 25.05 | 12.65 | 7.41 | 4.08 | 2.86 | 2.38 |

Table 33.7. Sol.27.7fg- Effort.

| England \& Wales |  |  |  |  |  |  | Belgium |  | Ireland |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Otter trawl ${ }^{1}$ | Beam trawl ${ }^{1}$ | Otter trawl ${ }^{2}$ | Beam trawl ${ }^{2}$ | Otter trawl ${ }^{3}$ | Beam trawl ${ }^{3}$ | Beam trawl ${ }^{4}$ | Beam trawl ${ }^{6}$ | Otter trawl ${ }^{5}$ | Scottish seine ${ }^{6}$ | Beam trawl ${ }^{6}$ |
| 1971 |  |  |  |  |  |  | 11.06 | - | - | - | - |
| 1972 |  |  |  |  |  |  | 8.44 | - | - | - | - |
| 1973 |  |  |  |  |  |  | 17.39 | - | - | - | - |
| 1974 |  |  |  |  |  |  | 18.83 | - | - | - | - |
| 1975 |  |  |  |  |  |  | 16.38 | - | - | - | - |
| 1976 |  |  |  |  |  |  | 28.07 | - | - | - | - |
| 1977 |  |  |  |  |  |  | 24.11 | - | - | - | - |
| 1978 |  |  |  |  |  |  | 18.09 | - | - | - | - |
| 1979 |  |  |  |  |  |  | 18.90 | - | - | - | - |
| 1980 |  |  |  |  |  |  | 29.02 | - | - | - | - |
| 1981 |  |  |  |  |  |  | 35.39 | - | - | - | - |
| 1982 |  |  |  |  |  |  | 28.77 | - | - | - | - |
| 1983 | 620 | 195 | 82 | 149 | 0 | 8 | 34.95 | - | - | - | - |
| 1984 | 1723 | 901 | 316 | 298 | 0 | 129 | 33.48 | - | - | - | - |
| 1985 | 1493 | 1101 | 206 | 285 | 23 | 92 | 40.49 | - | - | - | - |
| 1986 | 1125 | 973 | 334 | 180 | 35 | 29 | 52.46 | - | - | - | - |
| 1987 | 1211 | 1681 | 364 | 187 | 26 | 26 | 37.26 | - | - | - | - |
| 1988 | 838 | 1102 | 351 | 77 | 20 | 36 | 42.92 | - | - | - | - |
| 1989 | 966 | 861 | 327 | 125 | 15 | 7 | 53.58 | - | - | - | - |
| 1990 | 1229 | 1256 | 435 | 165 | 24 | 194 | 40.27 | - | - | - | - |
| 1991 | 1066 | 1667 | 306 | 483 | 45 | 104 | 18.05 | - | - | - | - |
| 1992 | 898 | 1420 | 303 | 633 | 435 | 90 | 25.47 | - | - | - | - |
| 1993 | 836 | 1669 | 251 | 694 | 30 | 135 | 31.27 | - | - | - | - |
| 1994 | 623 | 2219 | 225 | 610 | 19 | 116 | 38.35 | - | - | - | - |
| 1995 | 580 | 2303 | 196 | 694 | 30 | 128 | 47.81 | - | 63.33 | 6.43 | 20.69 |
| 1996 | 593 | 2391 | 341 | 560 | 105 | 220 | 47.63 | 53.27 | 59.97 | 9.73 | 26.70 |
| 1997 | 577 | 2661 | 370 | 770 | 122 | 146 | 51.98 | 57.36 | 65.00 | 16.07 | 28.06 |
| 1998 | 517 | 2846 | 385 | 591 | 94 | 159 | 52.11 | 57.79 | 72.25 | 14.88 | 35.21 |
| 1999 | 395 | 3058 | 176 | 1461 | 235 | 312 | 55.03 | 55.11 | 51.48 | 8.01 | 40.83 |
| 2000 | 284 | 3133 | 187 | 1007 | 160 | 200 | 56.05 | 51.34 | 60.56 | 9.86 | 36.83 |
| 2001 | 309 | 3172 | 187 | 1155 | 179 | 91 | 52.06 | 54.90 | 69.37 | 16.33 | 39.50 |
| 2002 | 416 | 2652 | 123 | 463 | 170 | 60 | 43.24 | 49.60 | 77.20 | 20.88 | 31.49 |
| 2003 | 696 | 2669 | 51 | 772 | 124 | 158 | 42.81 | 62.73 | 86.78 | 20.07 | 49.22 |
| 2004 | 641 | 2503 | 198 | 923 | 125 | 178 | - | 78.73 | 97.12 | 18.42 | 54.89 |
| 2005 | 876 | 1968 | 21 | 618 | 154 | 116 | - | 64.50 | 124.67 | 14.64 | 49.56 |
| 2006 | 924 | 1330 | 23 | 630 | 233 | 70 | - | 49.61 | 118.04 | 14.78 | 60.47 |
| 2007 | 798 | 1407 | 31 | 518 | 219 | 12 | - | 45.91 | 135.36 | 15.81 | 55.81 |
| 2008 | 711 | 1202 | 109 | 290 | 229 | 5 | - | 28.72 | 125.41 | 11.65 | 37.20 |
| 2009 | 656 | 1105 | 244 | 266 | 296 | 48 | - | 30.65 | 137.11 | 8.18 | 37.94 |
| 2010 | 565 | 1162 | 84 | 327 | 469 | 78 | - | 32.46 | 140.79 | 9.68 | 40.22 |
| 2011 | 525 | 868 | 8 | 180 | 353 | 111 | - | 38.77 | 120.33 | 11.01 | 35.33 |
| 2012 | 543 | 1408 | 138 | 275 | 487 | 102 | - | 46.25 | 127.68 | 14.14 | 40.33 |
| $\underline{2013}$ | 280 | 1611 | 72 | 265 | 37 | 77 | - | 45.23 | 118.20 | 13.15 | 38.48 |
| 2014 | 156 | 959 | 10 | 131 | 0 | 24 | - | 31.30 | 127.34 | 12.46 | 37.84 |
| $\underline{2015}$ | 79 | 726 | 3 | 245 | 0 | 56 | - | 31.79 | 132.69 | 9.28 | 37.79 |
| $\underline{2016}$ | 0 | 915 | 0 | 396 | 0 | 34 | - | 32.34 | 148.17 | 10.44 | 39.55 |
| 2017 | 93 | 986 | 95 | 514 | 193 | 74 | - | 33.35 | 136.05 | 9.75 | 35.21 |
| 2018 | 127 | 1071 | 71 | 440 | 210 | 15 | - | 31.48 | 105.81 | 9.69 | 37.42 |
| 2019 | 169 | 981 | 34 | 255 | 277 | 8 | - | 32.03 | 103.89 | 14.26 | 34.08 |
| 2020 | 100 | 1012 | 10 | 346 | 40 | 99 | - | 41.70 | 89.91 | 13.59 | 29.14 |
| 2021 | 155 | 1260 | 22 | 540 | 28 | 102 | - | 36.18 | 33.04 | 11.85 | 31.57 |

${ }^{1}$ Division 7.f only -days fished (Corrected).
${ }^{2} 7$.g EAST - days fished (corrected).
${ }^{3} 7 . \mathrm{g}$ WEST - days fished (corrected).
${ }^{4}$ Fishing hours ( $\times 10^{\wedge} 3$ ) corrected for fishing power using $P=0.000204$ BHP ${ }^{\wedge} 1.23$.
${ }^{5}$ Division 7.g only - Fishing hours (x10^3).
${ }^{6}$ Fishing hours ( $\mathbf{x 1 0 \wedge 3 \text { ). }}$

Table 33.8. Sol.27.7fg - LPUE.

|  | UK |  |  |  |  |  |  | Belgium |  | Ireland |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | BT | Otter | Beam | Otter | Beam | Otter | Beam | Beam | Beam | Otter | Scottish | Beam |
|  | Survey ${ }^{1}$ | ${ }^{1}$ trawl ${ }^{2}$ | trawl ${ }^{2}$ | trawl ${ }^{2}$ | trawl ${ }^{2}$ | trawl ${ }^{2}$ | trawl ${ }^{2}$ | trawl ${ }^{3}$ | trawl ${ }^{4}$ | trawl ${ }^{4}$ | sein ${ }^{4}$ | trawl ${ }^{4}$ |
| Year | Div | Div | Div | Div | Div | Div | Div | Div | Div | Div | Div | Div |
|  | VIIfg | VIIf | VIIf | VIIgEast | VIIgEast | VIIgWest | VIIgWest | VIIfg | VIIfg | VIIg | VIIg | VIIg |
| 1971 | - | - | - | - | - | - | - | 47.92 | - | - | - | - |
| 1972 | - | - | - | - | - | - | - | 37.06 | - | - | - | - |
| 1973 | - | - | - | - | - | - | - | 39.47 | - | - | - | - |
| 1974 | - | - | - | - | - | - | - | 37.81 | - | - | - | - |
| 1975 | - | - | - | - | - | - | - | 31.41 | - | - | - | - |
| 1976 | - | - | - | - | - | - | - | 30.50 | - | - | - | - |
| 1977 | - | - | - | - | - | - | - | 27.90 | - | - | - | - |
| 1978 | - | - | - | - | - | - | - | 23.35 | - | - | - | - |
| 1979 | - | - | - | - | - | - | - | 33.19 | - | - | - | - |
| 1980 | - | - | - | - | - | - | - | 29.73 | - | - | - | - |
| 1981 | - | - | - | - | - | - | - | 24.03 | - | - | - | - |
| 1982 | - | - | - | - | - | - | - | 25.93 | - | - | - | - |
| 1983 | - | 30.54 | 201.80 | 35.75 | 250.70 | 0.00 | 39.68 | 22.18 | - | - | - | - |
| 1984 | - | 19.53 | 204.65 | 28.04 | 130.61 | 0.00 | 63.21 | 20.78 | - | - | - | - |
| 1985 | - | 26.58 | 240.45 | 37.31 | 235.62 | 33.78 | 188.28 | 17.94 | - | - | - | - |
| 1986 | - | 25.55 | 247.74 | 21.27 | 190.11 | 10.22 | 184.94 | 17.83 | - | - | - | - |
| 1987 | - | 19.85 | 179.34 | 36.02 | 225.56 | 0.47 | 113.56 | 17.32 | - | - | - | - |
| 1988 | 79.52 | 11.13 | 110.35 | 8.88 | 304.43 | 1.82 | 230.65 | 15.29 | - | - | - | - |
| 1989 | 150.02 | 17.36 | 130.42 | 18.75 | 247.17 | 10.28 | 707.10 | 11.33 | - | - | - | - |
| 1990 | 93.61 | 13.41 | 148.47 | 18.08 | 269.40 | 8.12 | 106.57 | 15.64 | - | - | - | - |
| 1991 | 122.06 | 12.26 | 119.52 | 16.20 | 117.12 | 15.23 | 169.61 | 24.24 | - | - | - | - |
| 1992 | 121.41 | 17.90 | 105.84 | 20.99 | 119.32 | 20.62 | 127.52 | 18.57 | - | - | - | - |
| 1993 | 76.37 | 8.85 | 118.08 | 4.27 | 119.85 | 9.83 | 358.96 | 15.21 | - | - | - | - |
| 1994 | 109.74 | 13.00 | 70.00 | 3.50 | 74.32 | 5.72 | 116.30 | 13.94 | - | - | - | - |
| 1995 | 69.91 | 13.76 | 73.20 | 12.75 | 63.20 | 15.20 | 41.46 | 13.62 | - | 0.40 | 0.62 | 0.81 |
| 1996 | 71.71 | 9.69 | 65.05 | 6.95 | 43.84 | 0.68 | 12.41 | 11.27 | 11.45 | 0.73 | 0.05 | 0.88 |
| 1997 | 81.67 | 12.55 | 53.81 | 6.42 | 43.77 | 0.44 | 16.05 | 9.96 | 9.68 | 0.42 | 0.23 | 1.16 |
| 1998 | 137.11 | 8.24 | 44.86 | 4.85 | 27.16 | 0.04 | 47.84 | 10.12 | 9.64 | 0.48 | 0.11 | 1.11 |
| 1999 | 168.46 | 13.25 | 52.36 | 8.18 | 26.19 | 0.01 | 14.01 | 11.26 | 12.14 | 0.17 | 0.09 | 0.50 |
| 2000 | 228.46 | 7.01 | 53.85 | 23.26 | 36.94 | 0.09 | 14.9 | 11.90 | 13.77 | 0.19 | 0.05 | 0.26 |
| 2001 | 158.08 | 17.1 | 62.39 | 27.5 | 33.01 | 0.11 | 22.69 | 13.25 | 13.60 | 0.31 | 0.55 | 0.18 |
| 2002 | 121.89 | 11.61 | 79.47 | 47.01 | 54.15 | 0.11 | 43.04 | 18.71 | 17.80 | 0.43 | 0.29 | 0.14 |
| 2003 | 123.91 | 8.03 | 80.85 | 0.00 | 45.42 | 0.70 | 52.96 | 19.48 | 11.40 | 0.12 | 0.03 | 0.19 |
| 2004 | 152.03 | 8.84 | 76.09 | 2.70 | 37.88 | 0.05 | 91.33 | - | 9.17 | 0.19 | 0.02 | 0.20 |
| 2005 | 76.28 | 10.67 | 70.02 | 3.07 | 41.36 | 0.20 | 80.99 | - | 9.78 | 0.14 | 0.00 | 0.29 |
| 2006 | 68.96 | 16.40 | 81.57 | 6.23 | 45.13 | 0.10 | 20.93 | - | 10.63 | 0.11 | 0.05 | 0.26 |
| 2007 | 80.95 | 10.75 | 92.17 | 15.04 | 43.57 | 0.05 | 39.00 | - | 11.53 | 0.13 | 0.02 | 0.20 |
| 2008 | 115.96 | 11.94 | 94.85 | 10.67 | 41.48 | 0.00 | 19.96 | - | 14.35 | 0.12 | 0.02 | 0.29 |
| 2009 | 90.64 | 13.13 | 69.37 | 6.88 | 50.65 | 0.00 | 9.81 | - | 14.01 | 0.10 | 0.00 | 0.28 |
| 2010 | 109.55 | 13.59 | 79.90 | 8.63 | 53.69 | 0.00 | 44.89 | - | 16.68 | 0.13 | 0.01 | 0.20 |
| 2011 | 99.47 | 20.78 | 109.20 | 4.47 | 98.38 | 0.00 | 50.73 | - | 17.90 | 0.19 | 0.01 | 0.20 |
| 2012 | 101.45 | 24.10 | 80.16 | 5.17 | 53.43 | 0.00 | 42.43 | - | 17.01 | 0.15 | 0.01 | 0.48 |
| 2013 | 119.38 | 27.81 | 82.82 | 4.62 | 44.52 | 0.07 | 39.60 | - | 16.54 | 0.14 | 0.01 | 0.65 |
| 2014 | 86.75 | 6.19 | 107.25 | 11.56 | 42.11 | 0 | 18.57 | - | 21.30 | 0.12 | - | 0.34 |
| 2015 | 85.45 | 51.13 | 103.07 | 5.62 | 57.39 | 0 | 42.64 | - | 20.14 | 0.11 | - | 0.31 |
| 2016 | 113.55 | 0.00 | 113.16 | 0 | 33.65 | 0 | 34.17 | - | 16.25 | 0.10 | 0.01 | 0.20 |
| 2017 | 111.38 | 31.29 | 100.03 | 18.09 | 35.05 | 0.22 | 58.81 | - | 15.72 | 0.18 | 0.05 | 0.22 |
| 2018 | 206.44 | 36.37 | 119.89 | 4.86 | 47.74 | 0.15 | 52.26 | - | 18.09 | 0.18 | - | 0.27 |
| 2019 | 150.04 | 46.55 | 129.79 | 11.12 | 61.33 | 0.12 | 23.35 | - | 23.08 | 0.25 | 0.00 | 0.26 |
| 2020 | 111.72 | 51.82 | 168.07 | 5.58 | 117.30 | 0 | 92.58 | - | 25.74 | 0.31 | 0.02 | 0.93 |
| 2021* | 112.38 | 89.52 | 156.87 | 6.41 | 78.38 | 9.80 | 136.10 | - | 22.04 | 0.27 | 0.02 | 0.54 |

${ }^{1} \mathrm{Kg} / 100 \mathrm{~km}$.
${ }^{2} \mathrm{Kg} /$ day.
${ }^{3} \mathrm{Kg} / \mathrm{hr}$ corrected for fishing power using $\mathrm{P}=0.000204 \mathrm{BHP}^{\wedge} 1.23$.
${ }^{4} \mathrm{Kg} /$ hour.
*Provisional.

Table 33.9. Sol.27.7fg - Tuning series.

| BE-CBT_1971_1983 | Belgium Beam trawl (Biomass tuning index) |  |
| :---: | :---: | :---: |
| 1971 | 1983 |  |
| 1 | 10 | 0 |
| 1 | -1 |  |
| 1 | 45.319 |  |
| 1 | 33.193 |  |
| 1 | 35.906 |  |
| 1 | 35.915 |  |
| 1 | 29.286 |  |
| 1 | 27.369 |  |
| 1 | 25.677 |  |
| 1 | 23.971 |  |
| 1 | 32.663 |  |
| 1 | 28.343 |  |
| 1 | 23.326 |  |
| 1 | 26.083 |  |
| $1$ | $20.742$ |  |
| BE-CBT_1984_1996 | Belgium Beam trawl (Biomass tuning index) |  |
| 1984 | 1996 |  |
| 1 | 10 | $0$ |
| 1 | -1 |  |
| 1 | 19.788 |  |
| 1 | 20.556 |  |
| 1 | 19.824 |  |
| 1 | 18.996 |  |
| 1 | 15.129 |  |
| 1 | 12.805 |  |
| 1 | 16.620 |  |
| 1 | 23.442 |  |
| 1 | 20.455 |  |




| UK(E\&W)-CBT_2006_2021 | UK(E+W) Beam trawl (Biomass tuning index) |
| :---: | :---: |
| 2006 | 2021 |
| 1 | 10 |
| 1 | -1 |
| 1 | 131.32 |
| 1 | 218.97 |
| 1 | 241.55 |
| 1 | 208.66 |
| 1 | 233 |
| 1 | 281.56 |
| 1 | 248.56 |
| 1 | 211.97 |
| 1 | 295.78 |
| 1 | 235.16 |
| 1 | 201.64 |
| 1 | 205.73 |
| 1 | 254.71 |
| 1 | 233.35 |
| 1 | 292.91 |
| 1 | 357.47 |


| UK(E\&W)-BTS-Q3 | UK(E+W) 7.f Corystes (automated indices since 1995). Ages used in the assessment are in bold |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 2021 |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.75 | 0.85 |  |  |  |  |  |  |  |
| 0 | 9 |  |  |  |  |  |  |  |  |  |
| 74.120 | 22 | 60 | 242 | 36 | 14 | 4 | 0 | 0 | 0 | 0 |
| 91.909 | 132 | 204 | 304 | 162 | 18 | 14 | 6 | 4 | 2 | 2 |
| 69.858 | 21 | 269 | 219 | 35 | 11 | 3 | 5 | 2 | 0 | 0 |
| 123.410 | 40 | 297 | 638 | 83 | 21 | 18 | 5 | 0 | 3 | 2 |
| 125.078 | 5 | 493 | 325 | 174 | 37 | 23 | 12 | 1 | 2 | 1 |
| 127.672 | 6 | 207 | 436 | 52 | 28 | 3 | 2 | 2 | 1 | 1 |
| 120.816 | 1 | 424 | 430 | 133 | 23 | 11 | 9 | 0 | 0 | 3 |
| 114.886 | 31 | 142 | 255 | 60 | 13 | 7 | 14 | 1 | 1 | 1 |
| 118.592 | 3 | 178 | 251 | 64 | 27 | 7 | 3 | 4 | 1 | 3 |
| 114.886 | 37 | 498 | 207 | 21 | 13 | 14 | 5 | 3 | 6 | 0 |
| 114.886 | 104 | 885 | 472 | 58 | 11 | 9 | 5 | 2 | 1 | 5 |
| 118.592 | 29 | 2922 | 297 | 38 | 16 | 7 | 4 | 5 | 1 | 0 |
| 118.592 | 16 | 1086 | 1608 | 37 | 26 | 6 | 0 | 2 | 1 | 1 |
| 118.592 | 26 | 449 | 711 | 307 | 23 | 9 | 6 | 2 | 0 | 2 |
| 118.592 | 9 | 786 | 283 | 151 | 121 | 14 | 7 | 2 | 3 | 0 |
| 118.592 | 14 | 465 | 628 | 55 | 30 | 56 | 9 | 3 | 3 | 0 |
| 114.886 | 64 | 860 | 434 | 99 | 15 | 22 | 42 | 4 | 3 | 0 |
| 118.592 | 44 | 407 | 267 | 38 | 16 | 7 | 5 | 17 | 1 | 2 |
| 118.592 | 13 | 324 | 238 | 47 | 16 | 8 | 0 | 2 | 12 | 0 |
| 118.592 | 108 | 424 | 128 | 51 | 16 | 8 | 7 | 3 | 4 | 13 |
| 118.592 | 6 | 1232 | 124 | 15 | 18 | 7 | 9 | 4 | 3 | 5 |
| 118.592 | 1 | 604 | 377 | 29 | 8 | 10 | 4 | 3 | 3 | 2 |
| 118.592 | 21 | 101 | 558 | 144 | 20 | 2 | 5 | 9 | 4 | 2 |
| 118.592 | 21 | 595 | 62 | 164 | 82 | 8 | 2 | 7 | 3 | 0 |
| 118.592 | 16 | 643 | 274 | 9 | 63 | 28 | 1 | 1 | 1 | 3 |
| 118.592 | 11 | 331 | 614 | 51 | 16 | 29 | 18 | 1 | 6 | 1 |


| UK(E\&W)-BTS-Q3 | UK(E+W) 7.f Corystes (automated indices since 1995 ). Ages used in the assessment are in bold |  |  |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 118.592 | 40 | 289 | 305 | 90 | 16 | 6 | 27 | 9 | 1 | 1 |
| 118.592 | 33 | 885 | 57 | 52 | 37 | 8 | 4 | 16 | 7 | 0 |
| 118.592 | 31 | 680 | 426 | 14 | 32 | 15 | 8 | 4 | 6 | 9 |
| 118.592 | 7 | 1240 | 206 | 80 | 15 | 19 | 20 | 5 | 4 | 13 |
| 118.592 | 32 | 515 | 1074 | 331 | 53 | 20 | 11 | 18 | 13 | 5 |
| 118.592 | 3 | 840 | 386 | 195 | 27 | 34 | 4 | 7 | 8 | 5 |
| 118.592 | 68 | 1062 | 183 | 53 | 44 | 39 | 14 | 14 | 0 | 3 |
| 118.592 | 393 | 282 | 88 | 80 | 28 | 20 | 2 | 7 | 8 |  |

## Table 33.10. Sol.27.7fg - Configuration.

\$minAge
[1] 1
\$maxAge
[1] 10
\$maxAgePlusGroup
[1] 1000000
\$keyLogFsta
$[, 1][, 2][, 3][, 4][, 5][, 6][, 7][, 8][, 9][, 10]$



[4,] $-1 \begin{array}{llllllllll}1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$

[6,] $\quad-1 \begin{array}{llllllllll} & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$

\$corFlag
[1] 2
\$keyLogFpar
$[, 1][, 2][, 3][, 4][, 5][, 6][, 7][, 8][, 9][, 10]$
[1,] $\begin{array}{ccccccccccc}1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
[2,] $\begin{array}{lllllllllll}0 & 1 & 2 & 3 & 3 & -1 & -1 & -1 & -1 & -1\end{array}$

[4,] $\begin{array}{ccccccccccc}5 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$


[7,] $\quad 8 \quad-1$
\$keyQpow
$[, 1][, 2][, 3][, 4][, 5][, 6][, 7][, 8][, 9][, 10]$
[1,] $-1 \begin{array}{llllllllll} & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$

[3,] $\quad-1 \begin{array}{llllllllll} & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
[4,] $-1 \begin{array}{llllllllll} & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$

[6,] $\quad-1$-1
[7,] $\quad-1 \begin{array}{llllllllll} & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
\$keyVarF
$[, 1][, 2][, 3][, 4][, 5][, 6][, 7][, 8][, 9][, 10]$
[1,] $00 \begin{array}{llllllllll} & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$


[4,] $-1 \begin{array}{llllllllll} & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$

[6,] $\quad-1 \begin{array}{llllllllll} & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
[7,] $\begin{array}{ccccccccccc}1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
\$keyVarLogN
[1] 0111111111

## \$keyVarObs

[,1] [,2] [,3] [,4] [,5] [,6] [,7] [,8] [,9] [,10]
$[1] \quad 0 \quad 0 \quad 1 \quad 1 \quad 1 \quad 1 \quad 1 \quad 1 \quad 1 \quad$,
[2,] $22 \begin{array}{llllllllll} & 3 & 3 & 4 & 4 & -1 & -1 & -1 & -1 & -1\end{array}$

[4,] $6 \begin{array}{llllllllll} & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
[5,] 7 7 -1 -1 $-1 \begin{array}{lllllll} & -1 & -1 & -1 & -1 & -1 & -1\end{array}$


\$obsCorStruct
[1] AR ID ID ID ID ID ID
Levels: ID AR US
\$keyCorObs
1-2 2-3 3-4 4-5 5-6 6-7 7-8 8-9 9-10
[1,] 0000000000
[2,] NA NA NA NA -1 -1 -1 -1 -1
[3,] $-1 \begin{array}{llllllll}1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
[4,] -1 -1
[5,] -1 -1
[6,] -1 -1
[7,] $-1 \begin{array}{llllllll}1 & -1 & -1 & -1 & -1 & -1 & -1 & -1\end{array}$
\$stockRecruitmentModelCode
[1] 0
\$noScaledYears
[1] 0
\$keyScaledYears
numeric(0)
\$keyParScaledYA
$<0 \times 0$ matrix>
\$fbarRange
[1] 38
\$keyBiomassTreat
[1]-1-1 000000
\$obsLikelihoodFlag
[1] LN LN LN LN LN LN LN
Levels: LN ALN
\$fixVarToWeight
[1] 0
\$fracMixF
[1] 0
\$fracMixN
[1] 0
\$fracMixObs
[1] 0000000

```
$constRecBreaks
numeric(0)
$predVarObsLink
    [,1] [,2] [,3] [,4] [,5] [,6] [,7] [,8] [,9] [,10]
[1,] 
[2,] -1 -1 -1 -1 -1 -1 NA NA NA NA NA
[3,] NA NA NA NA NA NA NA NA NA NA
[4,] NA NA NA NA NA NA NA NA NA NA
[5,] NA NA NA NA NA NA NA NA NA NA
[6,] NA NA NA NA NA NA NA NA NA NA
[7,] NA NA NA NA NA NA NA NA NA NA
$hockeyStickCurve
[1] }2
$stockWeightModel
[1] 0
$keyStockWeightMean
[1] NA NA NA NA NA NA NA NA NA NA
$keyStockWeightObsVar
[1] NA NA NA NA NA NA NA NA NA NA
$catchWeightModel
[1] 0
$keyCatchWeightMean
    [1] NA NA NA NA NA NA NA NA NA NA
$keyCatchWeightObsVar
[1] NA NA NA NA NA NA NA NA NA NA
$matureModel
[1] 0
$keyMatureMean
[1] NA NA NA NA NA NA NA NA NA NA
$mortalityModel
[1] 0
$keyMortalityMean
[1] NA NA NA NA NA NA NA NA NA NA
$keyMortalityObsVar
[1] NA NA NA NA NA NA NA NA NA NA
$keyXtraSd
    [,1] [,2] [,3] [,4]
```

Table 33.11. Sol.27.7fg - Diagnostics.

```
#Model_table#
    log(L) #par AIC
M1 -249.3667 24 546.7334
```

\#Tuning_table\#
Name Type Years Ages LogQ_age1 sd_age1 LogQ_age2 sd_age2 LogQ_age3 sd_age3
UK-BTS-Q3 age-based 1988-2021 $1-5 \quad-7.196 \quad 0.100 \quad 10-7.4 \quad 0.1 \quad-8.614 \quad 0.099$
BE-CBT_71-83 biomass 1971-1983 -1 -5.115 0.064 NA NA NA NA
BE-CBT_84-96 biomass 1984-1996 -1
BE-CBT_06-21 biomass 2006-2021 -1 -7.360 NA NA NA NA

| UK-CBT_-84-05 | biomass | $1984-2005$ | -1 | -3.825 | 0.076 | NA | NA | NA |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| UK-CBT_06-21 | biomass | $2006-2021$ | -1 | -2.804 | 0.074 | NA | NA | NA | NA

Name Type Years Ages LogQ_age4 sd_age4 LogQ_age5 sd_age5 $\begin{array}{lllllll}\text { UK-BTS-Q3 age-based } & 1988-2021 & 1-5 & -9.239 & -0.066 & -9.239 & -0.066\end{array}$ BE-CBT_71-83 biomass 1971-1983 -1 NA NA NA NA
BE-CBT_84-96 biomass 1984-1996 -1
BE-CBT_06-21 biomass 2006-2021 -1
UK-CBT_84-05 biomass 1984-2005 -1 NA NA NA NA
UK-CBT_06-21 biomass 2006-2021 -1 NA NA NA NA

Table 33.12. Sol.27.7fg - Fishing mortality.

| year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Fbar(3-8) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0.003 | 0.086 | 0.281 | 0.33 | 0.33 | 0.33 | 0.293 | 0.293 | 0.273 | 0.273 | 0.31 |
| 1972 | 0.003 | 0.08 | 0.261 | 0.304 | 0.304 | 0.304 | 0.268 | 0.268 | 0.249 | 0.249 | 0.285 |
| 1973 | 0.003 | 0.076 | 0.246 | 0.282 | 0.282 | 0.282 | 0.248 | 0.248 | 0.231 | 0.231 | 0.265 |
| 1974 | 0.003 | 0.072 | 0.233 | 0.267 | 0.267 | 0.267 | 0.236 | 0.236 | 0.22 | 0.22 | 0.251 |
| 1975 | 0.003 | 0.07 | 0.228 | 0.26 | 0.26 | 0.26 | 0.23 | 0.23 | 0.213 | 0.213 | 0.244 |
| 1976 | 0.003 | 0.081 | 0.264 | 0.299 | 0.299 | 0.299 | 0.264 | 0.264 | 0.243 | 0.243 | 0.282 |
| 1977 | 0.003 | 0.074 | 0.24 | 0.271 | 0.271 | 0.271 | 0.238 | 0.238 | 0.218 | 0.218 | 0.255 |
| 1978 | 0.003 | 0.068 | 0.219 | 0.247 | 0.247 | 0.247 | 0.215 | 0.215 | 0.196 | 0.196 | 0.231 |
| 1979 | 0.003 | 0.072 | 0.233 | 0.263 | 0.263 | 0.263 | 0.228 | 0.228 | 0.207 | 0.207 | 0.247 |
| 1980 | 0.003 | 0.082 | 0.265 | 0.3 | 0.3 | 0.3 | 0.259 | 0.259 | 0.233 | 0.233 | 0.281 |
| 1981 | 0.003 | 0.092 | 0.299 | 0.339 | 0.339 | 0.339 | 0.293 | 0.293 | 0.262 | 0.262 | 0.317 |
| 1982 | 0.004 | 0.096 | 0.314 | 0.358 | 0.358 | 0.358 | 0.311 | 0.311 | 0.276 | 0.276 | 0.335 |
| 1983 | 0.004 | 0.101 | 0.33 | 0.378 | 0.378 | 0.378 | 0.329 | 0.329 | 0.289 | 0.289 | 0.354 |
| 1984 | 0.004 | 0.101 | 0.332 | 0.385 | 0.385 | 0.385 | 0.332 | 0.332 | 0.291 | 0.291 | 0.358 |
| 1985 | 0.004 | 0.108 | 0.355 | 0.417 | 0.417 | 0.417 | 0.357 | 0.357 | 0.313 | 0.313 | 0.386 |
| 1986 | 0.004 | 0.117 | 0.385 | 0.46 | 0.46 | 0.46 | 0.393 | 0.393 | 0.345 | 0.345 | 0.425 |
| 1987 | 0.004 | 0.114 | 0.375 | 0.455 | 0.455 | 0.455 | 0.389 | 0.389 | 0.341 | 0.341 | 0.42 |
| 1988 | 0.004 | 0.11 | 0.363 | 0.447 | 0.447 | 0.447 | 0.38 | 0.38 | 0.334 | 0.334 | 0.411 |
| 1989 | 0.004 | 0.105 | 0.344 | 0.424 | 0.424 | 0.424 | 0.358 | 0.358 | 0.313 | 0.313 | 0.389 |
| 1990 | 0.004 | 0.11 | 0.361 | 0.447 | 0.447 | 0.447 | 0.375 | 0.375 | 0.329 | 0.329 | 0.408 |
| 1991 | 0.004 | 0.106 | 0.348 | 0.432 | 0.432 | 0.432 | 0.36 | 0.36 | 0.32 | 0.32 | 0.394 |
| 1992 | 0.004 | 0.104 | 0.342 | 0.423 | 0.423 | 0.423 | 0.35 | 0.35 | 0.314 | 0.314 | 0.385 |
| 1993 | 0.004 | 0.111 | 0.366 | 0.455 | 0.455 | 0.455 | 0.378 | 0.378 | 0.339 | 0.339 | 0.414 |
| 1994 | 0.005 | 0.12 | 0.395 | 0.495 | 0.495 | 0.495 | 0.41 | 0.41 | 0.37 | 0.37 | 0.45 |
| 1995 | 0.005 | 0.132 | 0.435 | 0.546 | 0.546 | 0.546 | 0.451 | 0.451 | 0.406 | 0.406 | 0.496 |
| 1996 | 0.005 | 0.133 | 0.44 | 0.551 | 0.551 | 0.551 | 0.455 | 0.455 | 0.41 | 0.41 | 0.501 |
| 1997 | 0.005 | 0.136 | 0.449 | 0.561 | 0.561 | 0.561 | 0.465 | 0.465 | 0.42 | 0.42 | 0.51 |
| 1998 | 0.005 | 0.128 | 0.421 | 0.526 | 0.526 | 0.526 | 0.436 | 0.436 | 0.395 | 0.395 | 0.479 |
| 1999 | 0.004 | 0.115 | 0.374 | 0.465 | 0.465 | 0.465 | 0.383 | 0.383 | 0.348 | 0.348 | 0.422 |
| 2000 | 0.004 | 0.1 | 0.324 | 0.403 | 0.403 | 0.403 | 0.334 | 0.334 | 0.306 | 0.306 | 0.367 |
| 2001 | 0.004 | 0.097 | 0.315 | 0.396 | 0.396 | 0.396 | 0.329 | 0.329 | 0.305 | 0.305 | 0.36 |
| 2002 | 0.004 | 0.099 | 0.319 | 0.406 | 0.406 | 0.406 | 0.338 | 0.338 | 0.315 | 0.315 | 0.369 |
| 2003 | 0.004 | 0.106 | 0.345 | 0.445 | 0.445 | 0.445 | 0.372 | 0.372 | 0.349 | 0.349 | 0.404 |
| $\underline{2004}$ | 0.004 | 0.109 | 0.355 | 0.46 | 0.46 | 0.46 | 0.387 | 0.387 | 0.367 | 0.367 | 0.418 |
| 2005 | 0.004 | 0.105 | 0.34 | 0.445 | 0.445 | 0.445 | 0.377 | 0.377 | 0.362 | 0.362 | 0.405 |
| $\underline{2006}$ | 0.004 | 0.1 | 0.321 | 0.424 | 0.424 | 0.424 | 0.364 | 0.364 | 0.354 | 0.354 | 0.387 |
| 2007 | 0.004 | 0.094 | 0.299 | 0.4 | 0.4 | 0.4 | 0.349 | 0.349 | 0.341 | 0.341 | 0.366 |
| 2008 | 0.003 | 0.086 | 0.271 | 0.367 | 0.367 | 0.367 | 0.326 | 0.326 | 0.321 | 0.321 | 0.337 |
| $\underline{2009}$ | 0.003 | 0.081 | 0.254 | 0.351 | 0.351 | 0.351 | 0.315 | 0.315 | 0.312 | 0.312 | 0.323 |
| 2010 | 0.003 | 0.077 | 0.242 | 0.338 | 0.338 | 0.338 | 0.306 | 0.306 | 0.303 | 0.303 | 0.311 |
| 2011 | 0.003 | 0.078 | 0.243 | 0.344 | 0.344 | 0.344 | 0.315 | 0.315 | 0.314 | 0.314 | 0.317 |
| $\underline{2012}$ | 0.003 | 0.082 | 0.256 | 0.368 | 0.368 | 0.368 | 0.34 | 0.34 | 0.341 | 0.341 | 0.34 |
| 2013 | 0.003 | 0.082 | 0.256 | 0.372 | 0.372 | 0.372 | 0.345 | 0.345 | 0.349 | 0.349 | 0.344 |
| 2014 | 0.003 | 0.081 | 0.253 | 0.37 | 0.37 | 0.37 | 0.344 | 0.344 | 0.352 | 0.352 | 0.342 |
| $\underline{2015}$ | 0.003 | 0.074 | 0.229 | 0.334 | 0.334 | 0.334 | 0.31 | 0.31 | 0.318 | 0.318 | 0.309 |
| 2016 | 0.003 | 0.067 | 0.206 | 0.301 | 0.301 | 0.301 | 0.277 | 0.277 | 0.284 | 0.284 | 0.277 |
| $\underline{2017}$ | 0.002 | 0.06 | 0.185 | 0.271 | 0.271 | 0.271 | 0.248 | 0.248 | 0.255 | 0.255 | 0.249 |
| $\underline{2018}$ | 0.002 | 0.056 | 0.172 | 0.253 | 0.253 | 0.253 | 0.232 | 0.232 | 0.239 | 0.239 | 0.232 |
| 2019 | 0.002 | 0.057 | 0.175 | 0.258 | 0.258 | 0.258 | 0.236 | 0.236 | 0.243 | 0.243 | 0.237 |
| 2020 | 0.002 | 0.06 | 0.185 | 0.274 | 0.274 | 0.274 | 0.251 | 0.251 | 0.256 | 0.256 | 0.251 |
| $\underline{2021}$ | 0.002 | 0.06 | 0.184 | 0.274 | 0.274 | 0.274 | 0.249 | 0.249 | 0.254 | 0.254 | 0.25 |

Table 33.13. Sol.27.7fg - Stock numbers-at-age (start of year, in thousands).

| year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 8764 | 5281 | 1833 | 4682 | 2034 | 1514 | 1804 | 2652 | 1462 | 4827 | 34853 |
| 1972 | 4526 | 8318 | 4355 | 1400 | 2805 | 1265 | 999 | 1153 | 1722 | 4009 | 30552 |
| 1973 | 3408 | 3965 | 7335 | 2940 | 1085 | 1744 | 786 | 711 | 766 | 3815 | 26556 |
| 1974 | 3476 | 3003 | 3228 | 5032 | 2012 | 794 | 1144 | 545 | 529 | 3073 | 22835 |
| 1975 | 3507 | 3139 | 2460 | 2325 | 3445 | 1407 | 548 | 773 | 405 | 2408 | 20418 |
| 1976 | 5278 | 3068 | 2653 | 1840 | 1748 | 2390 | 979 | 368 | 602 | 1976 | 20903 |
| 1977 | 4852 | 4946 | 2482 | 1664 | 1325 | 1135 | 1755 | 673 | 220 | 1800 | 20851 |
| 1978 | 5208 | 4325 | 4327 | 1718 | 1041 | 1001 | 776 | 1278 | 480 | 1324 | 21478 |
| 1979 | 3715 | 4851 | 3648 | 3300 | 1157 | 729 | 781 | 608 | 961 | 1301 | 21050 |
| 1980 | 4903 | 3196 | 4162 | 2669 | 2173 | 809 | 507 | 556 | 465 | 1694 | 21136 |
| 1981 | 5055 | 4508 | 2540 | 2917 | 1687 | 1364 | 525 | 418 | 387 | 1561 | 20961 |
| 1982 | 5038 | 4570 | 3815 | 1550 | 1937 | 1094 | 832 | 334 | 326 | 1382 | 20877 |
| 1983 | 6467 | 4453 | 3750 | 2476 | 977 | 1242 | 701 | 522 | 213 | 1254 | 22056 |
| 1984 | 5698 | 6002 | 3622 | 2374 | 1592 | 611 | 837 | 456 | 298 | 1103 | 22591 |
| 1985 | 5600 | 5093 | 5059 | 2308 | 1530 | 950 | 399 | 536 | 302 | 989 | 22765 |
| 1986 | 3947 | 5178 | 4094 | 3169 | 1376 | 902 | 555 | 275 | 337 | 904 | 20736 |
| 1987 | 5677 | 3365 | 4274 | 2404 | 1799 | 778 | 499 | 325 | 186 | 785 | 20092 |
| 1988 | 4605 | 5347 | 2567 | 2837 | 1277 | 1043 | 453 | 267 | 204 | 674 | 19274 |
| 1989 | 4856 | 4327 | 4518 | 1660 | 1732 | 700 | 591 | 295 | 148 | 566 | 19394 |
| 1990 | 7394 | 4334 | 3659 | 2982 | 968 | 1069 | 404 | 356 | 189 | 504 | 21859 |
| 1991 | 4950 | 7192 | 3420 | 2366 | 1717 | 548 | 618 | 242 | 207 | 492 | 21751 |
| 1992 | 5045 | 4387 | 6100 | 2112 | 1382 | 953 | 321 | 379 | 151 | 442 | 21272 |
| 1993 | 4728 | 4550 | 3482 | 3748 | 1170 | 769 | 533 | 210 | 261 | 383 | 19833 |
| 1994 | 4313 | 4407 | 3643 | 2134 | 2128 | 674 | 461 | 296 | 136 | 428 | 18620 |
| 1995 | 3781 | 3868 | 3592 | 2263 | 1120 | 1155 | 362 | 279 | 184 | 350 | 16953 |
| 1996 | 4383 | 3447 | 3038 | 2139 | 1131 | 593 | 586 | 204 | 157 | 331 | 16009 |
| 1997 | 6338 | 4005 | 2757 | 1702 | 1091 | 586 | 314 | 323 | 120 | 297 | 17532 |
| 1998 | 7907 | 5773 | 3221 | 1588 | 869 | 541 | 295 | 176 | 177 | 251 | 20798 |
| 1999 | 13679 | 6834 | 4553 | 1964 | 772 | 463 | 296 | 155 | 107 | 246 | 29069 |
| $\underline{2000}$ | 9156 | 12696 | 5333 | 2637 | 1111 | 390 | 272 | 180 | 94 | 238 | 32105 |
| 2001 | 5702 | 8160 | 10410 | 3245 | 1520 | 687 | 248 | 179 | 112 | 243 | 30505 |
| 2002 | 7121 | 4880 | 6588 | 7204 | 1890 | 925 | 405 | 169 | 126 | 237 | 29544 |
| 2003 | 5429 | 6579 | 3820 | 4281 | 4494 | 1030 | 589 | 267 | 107 | 267 | 26863 |
| 2004 | 4806 | 4826 | 5398 | 2450 | 2638 | 2387 | 544 | 360 | 171 | 264 | 23844 |
| 2005 | 4283 | 4348 | 3782 | 3218 | 1440 | 1493 | 1296 | 322 | 219 | 288 | 20690 |
| 2006 | 3544 | 3895 | 3542 | 2324 | 1907 | 814 | 870 | 770 | 205 | 328 | 18199 |
| 2007 | 3924 | 3174 | 3214 | 2259 | 1356 | 1167 | 487 | 547 | 513 | 338 | 16978 |
| 2008 | 7507 | 3362 | 2574 | 2096 | 1369 | 821 | 729 | 302 | 368 | 558 | 19684 |
| 2009 | 6791 | 7111 | 2692 | 1811 | 1335 | 847 | 540 | 470 | 192 | 634 | 22423 |
| 2010 | 3110 | 6354 | 6602 | 2040 | 1197 | 873 | 552 | 362 | 321 | 552 | 21963 |
| 2011 | 5364 | 2554 | 5578 | 5161 | 1382 | 867 | 578 | 387 | 234 | 637 | 22741 |
| 2012 | 6362 | 4885 | 2017 | 3977 | 3424 | 959 | 587 | 380 | 257 | 623 | 23472 |
| 2013 | 5111 | 5780 | 4044 | 1535 | 2435 | 1986 | 637 | 379 | 240 | 584 | 22729 |
| $\underline{2014}$ | 3987 | 4719 | 4791 | 2843 | 1021 | 1456 | 1186 | 406 | 250 | 550 | 21209 |
| 2015 | 7926 | 3311 | 3989 | 3503 | 1748 | 653 | 904 | 670 | 268 | 467 | 23438 |
| 2016 | 8365 | 7404 | 2693 | 2757 | 2269 | 1130 | 447 | 565 | 440 | 429 | 26499 |
| $\underline{2017}$ | 14305 | 7059 | 6505 | 2075 | 1821 | 1455 | 760 | 320 | 362 | 564 | 35225 |
| 2018 | 6861 | 14068 | 5896 | 5100 | 1522 | 1244 | 934 | 548 | 241 | 617 | 37031 |
| 2019 | 7032 | 6159 | 12982 | 4469 | 3671 | 1212 | 825 | 625 | 432 | 599 | 38006 |
| $\underline{2020}$ | 6104 | 6079 | 5391 | 10338 | 3714 | 2536 | 925 | 570 | 442 | 764 | 36864 |
| 2021 | 8071 | 5392 | 5079 | 4396 | 7427 | 2946 | 1730 | 661 | 379 | 873 | 36952 |

Table 33.14. Sol.27.7fg - Summary ('Catch' refers to model estimate).

| year | $\begin{aligned} & \hline \mathrm{R} \\ & \text { low } \\ & \hline \end{aligned}$ | R value | $\begin{aligned} & \hline \mathrm{R} \\ & \text { high } \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { TSB } \\ & \text { low } \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { TSB } \\ & \text { value } \end{aligned}$ | $\begin{aligned} & \hline \text { TSB } \\ & \text { high } \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { SSB } \\ & \text { low } \\ & \hline \end{aligned}$ | SSB <br> value | $\begin{aligned} & \hline \text { SSB } \\ & \text { high } \end{aligned}$ | Catch low | Catch value | Catch high | $\begin{aligned} & \text { Fbar } \\ & \text { low } \\ & \hline \end{aligned}$ | Fbar value | Fbar <br> high |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 5929 | 8764 | 12954 | 6180 | 7196 | 8379 | 5642 | 6636 | 7805 | 1376 | 1699 | 2099 | 0.248 | 0.31 | 0.387 |
| 1972 | 3178 | 4526 | 6445 | 5356 | 6220 | 7224 | 4926 | 5762 | 6741 | 1185 | 1401 | 1658 | 0.234 | 0.285 | 0.346 |
| 1973 | 2392 | 3408 | 4855 | 5456 | 6316 | 7312 | 5062 | 5891 | 6855 | 1194 | 1409 | 1662 | 0.218 | 0.265 | 0.321 |
| 1974 | 2439 | 3476 | 4953 | 5013 | 5809 | 6731 | 4709 | 5484 | 6387 | 1049 | 1241 | 1470 | 0.206 | 0.251 | 0.306 |
| 1975 | 2452 | 3507 | 5015 | 4621 | 5355 | 6205 | 4338 | 5055 | 5891 | 957 | 1133 | 1342 | 0.2 | 0.244 | 0.299 |
| 1976 | 3726 | 5278 | 7475 | 4266 | 4933 | 5705 | 3925 | 4571 | 5323 | 912 | 1092 | 1307 | 0.23 | 0.282 | 0.345 |
| 1977 | 3432 | 4852 | 6859 | 4202 | 4831 | 5554 | 3770 | 4361 | 5045 | 825 | 970 | 1140 | 0.21 | 0.255 | 0.309 |
| 1978 | 3660 | 5208 | 7409 | 4266 | 4902 | 5632 | 3833 | 4431 | 5123 | 766 | 919 | 1102 | 0.187 | 0.231 | 0.287 |
| 1979 | 2589 | 3715 | 5331 | 4343 | 4978 | 5705 | 3986 | 4593 | 5292 | 863 | 1025 | 1217 | 0.202 | 0.247 | 0.301 |
| 1980 | 3470 | 4903 | 6928 | 4319 | 4943 | 5657 | 3990 | 4594 | 5288 | 1001 | 1181 | 1394 | 0.234 | 0.281 | 0.336 |
| 1981 | 3577 | 5055 | 7143 | 4051 | 4612 | 5251 | 3681 | 4217 | 4832 | 957 | 1133 | 1341 | 0.266 | 0.317 | 0.378 |
| 1982 | 3561 | 5038 | 7127 | 4049 | 4581 | 5182 | 3633 | 4134 | 4704 | 1039 | 1226 | 1447 | 0.281 | 0.335 | 0.398 |
| 1983 | 4556 | 6467 | 9178 | 4047 | 4564 | 5147 | 3589 | 4070 | 4616 | 1061 | 1254 | 1481 | 0.299 | 0.354 | 0.419 |
| 1984 | 4010 | 5698 | 8097 | 4214 | 4741 | 5334 | 3737 | 4219 | 4762 | 1086 | 1280 | 1509 | 0.303 | 0.358 | 0.423 |
| 1985 | 3941 | 5600 | 7957 | 3857 | 4341 | 4886 | 3468 | 3920 | 4431 | 1141 | 1351 | 1601 | 0.328 | 0.386 | 0.455 |
| 1986 | 2778 | 3947 | 5608 | 3992 | 4500 | 5073 | 3591 | 4058 | 4585 | 1217 | 1452 | 1733 | 0.359 | 0.425 | 0.504 |
| 1987 | 4043 | 5677 | 7971 | 3552 | 3998 | 4499 | 3173 | 3587 | 4055 | 1110 | 1315 | 1559 | 0.356 | 0.42 | 0.495 |
| 1988 | 3312 | 4605 | 6404 | 3288 | 3693 | 4147 | 2919 | 3293 | 3715 | 974 | 1155 | 1370 | 0.349 | 0.411 | 0.484 |
| 1989 | 3494 | 4856 | 6748 | 3146 | 3522 | 3942 | 2765 | 3108 | 3493 | 841 | 998 | 1185 | 0.33 | 0.389 | 0.459 |
| 1990 | 5296 | 7394 | 10324 | 2971 | 3324 | 3719 | 2570 | 2890 | 3250 | 934 | 1106 | 1309 | 0.345 | 0.408 | 0.484 |
| 1991 | 3583 | 4950 | 6837 | 3248 | 3669 | 4145 | 2826 | 3198 | 3619 | 947 | 1132 | 1353 | 0.333 | 0.394 | 0.466 |
| 1992 | 3671 | 5045 | 6933 | 3280 | 3708 | 4193 | 2877 | 3264 | 3703 | 891 | 1079 | 1307 | 0.323 | 0.385 | 0.46 |
| 1993 | 3429 | 4728 | 6518 | 3090 | 3469 | 3895 | 2709 | 3053 | 3439 | 906 | 1082 | 1291 | 0.351 | 0.414 | 0.488 |
| 1994 | 3137 | 4313 | 5931 | 2859 | 3200 | 3581 | 2512 | 2820 | 3167 | 914 | 1085 | 1288 | 0.385 | 0.45 | 0.526 |
| 1995 | 2731 | 3781 | 5236 | 2774 | 3100 | 3465 | 2443 | 2736 | 3065 | 963 | 1145 | 1362 | 0.423 | 0.496 | 0.582 |
| 1996 | 3166 | 4383 | 6068 | 2524 | 2818 | 3146 | 2207 | 2472 | 2768 | 883 | 1043 | 1231 | 0.428 | 0.501 | 0.585 |
| 1997 | 4588 | 6338 | 8754 | 2312 | 2589 | 2900 | 1948 | 2191 | 2465 | 831 | 988 | 1175 | 0.432 | 0.51 | 0.602 |
| 1998 | 5663 | 7907 | 11041 | 2652 | 2998 | 3391 | 2131 | 2409 | 2723 | 807 | 963 | 1151 | 0.405 | 0.479 | 0.566 |
| 1999 | 9744 | 13679 | 19202 | 3117 | 3543 | 4028 | 2352 | 2672 | 3037 | 779 | 938 | 1130 | 0.357 | 0.422 | 0.5 |
| 2000 | 6648 | 9156 | 12610 | 3157 | 3583 | 4065 | 2558 | 2909 | 3309 | 885 | 1091 | 1345 | 0.304 | 0.367 | 0.443 |
| 2001 | 4109 | 5702 | 7914 | 4405 | 5055 | 5801 | 3780 | 4349 | 5003 | 1121 | 1368 | 1670 | 0.302 | 0.36 | 0.43 |
| 2002 | 5166 | 7121 | 9816 | 4402 | 5022 | 5728 | 3871 | 4441 | 5094 | 1221 | 1477 | 1787 | 0.312 | 0.369 | 0.436 |
| 2003 | 3954 | 5429 | 7453 | 4136 | 4686 | 5309 | 3668 | 4176 | 4753 | 1243 | 1498 | 1805 | 0.343 | 0.404 | 0.475 |
| 2004 | 3525 | 4806 | 6553 | 4113 | 4631 | 5214 | 3636 | 4104 | 4632 | 1215 | 1452 | 1735 | 0.354 | 0.418 | 0.493 |
| 2005 | 3160 | 4283 | 5804 | 3720 | 4167 | 4667 | 3318 | 3729 | 4191 | 1104 | 1308 | 1550 | 0.343 | 0.405 | 0.477 |
| 2006 | 2607 | 3544 | 4818 | 3283 | 3666 | 4093 | 2936 | 3286 | 3678 | 873 | 1026 | 1206 | 0.328 | 0.387 | 0.456 |
| 2007 | 2886 | 3924 | 5336 | 3051 | 3415 | 3823 | 2700 | 3028 | 3397 | 819 | 960 | 1125 | 0.309 | 0.366 | 0.433 |
| 2008 | 5411 | 7507 | 10415 | 3129 | 3534 | 3992 | 2645 | 2987 | 3373 | 731 | 856 | 1001 | 0.283 | 0.337 | 0.402 |
| 2009 | 4982 | 6791 | 9256 | 3321 | 3814 | 4381 | 2724 | 3119 | 3572 | 623 | 734 | 865 | 0.271 | 0.323 | 0.384 |
| 2010 | 2236 | 3110 | 4326 | 3428 | 3878 | 4388 | 2982 | 3378 | 3827 | 716 | 844 | 994 | 0.261 | 0.311 | 0.37 |
| 2011 | 3953 | 5364 | 7280 | 3614 | 4066 | 4574 | 3211 | 3627 | 4098 | 867 | 1022 | 1204 | 0.267 | 0.317 | 0.377 |
| 2012 | 4690 | 6362 | 8630 | 3629 | 4085 | 4599 | 3139 | 3551 | 4017 | 849 | 1007 | 1195 | 0.284 | 0.34 | 0.406 |
| 2013 | 3759 | 5111 | 6950 | 3559 | 4002 | 4500 | 3078 | 3470 | 3912 | 848 | 999 | 1178 | 0.285 | 0.344 | 0.414 |
| 2014 | 2898 | 3987 | 5486 | 3639 | 4119 | 4661 | 3236 | 3674 | 4171 | 931 | 1100 | 1298 | 0.28 | 0.342 | 0.417 |
| 2015 | 5799 | 7926 | 10833 | 3584 | 4060 | 4599 | 3070 | 3503 | 3998 | 781 | 921 | 1086 | 0.252 | 0.309 | 0.377 |
| 2016 | 6197 | 8365 | 11292 | 3837 | 4360 | 4953 | 3152 | 3591 | 4092 | 688 | 813 | 959 | 0.225 | 0.277 | 0.34 |
| 2017 | 10348 | 14305 | 19774 | 4385 | 4990 | 5677 | 3489 | 3979 | 4539 | 699 | 826 | 976 | 0.201 | 0.249 | 0.308 |
| 2018 | 4956 | 6861 | 9498 | 5285 | 6036 | 6894 | 4415 | 5040 | 5754 | 804 | 960 | 1148 | 0.186 | 0.232 | 0.29 |
| 2019 | 4886 | 7032 | 10120 | 5759 | 6569 | 7494 | 5107 | 5846 | 6692 | 996 | 1182 | 1402 | 0.19 | 0.237 | 0.296 |
| 2020 | 3542 | 6104 | 10519 | 5653 | 6534 | 7552 | 5126 | 5945 | 6894 | 1122 | 1350 | 1624 | 0.197 | 0.251 | 0.321 |
| $\underline{2021}$ | 3880 | 8071 | 16786 | 5437 | 6579 | 7961 | 4879 | 5894 | 7119 | 1080 | 1317 | 1606 | 0.188 | 0.25 | 0.333 |

Table 33.15. Sol.27.7fg - Short-term forecast.

Assumptions made for the interim year and in the forecast.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| Fages 3-8 (2022) | 0.25 | Based on a catch of 1337 t for 2022. |
| SSB $_{2023}$ | 6120 | Short-term forecast fishing at F=0.25; in tonnes. |
| Rage $_{1}$ (2022-2023) | 5111 | Median recruitment, resampled from the years 1971-2019; in thou- <br> sands. |
| Catch (2022) | 1337 | TAC for 2022; in tonnes. |
| Projected landings (2022) | 1258 | Assuming average landings ratio by age 2019-2021; in tonnes. |
| Projected discards (2022) | 79 | Assuming average discard ratio by age 2019-2021; in tonnes. |

Annual catch scenarios. All weights are in tonnes.

| Basis | Total catch (2023) | Projected landings * (2023) | Projected discards <br> (2023) | $F_{\text {total }}$ (2023) | $F_{\text {pro- }}$ <br> jected <br> landings <br> (2023) | $F_{\text {pro- }}$ <br> jected discards (2023) | $\begin{aligned} & \text { SSB } \\ & (2024) \end{aligned}$ | $\begin{aligned} & \text { \% SSB } \\ & \text { change } \\ & * * * \end{aligned}$ | \% TAC <br> change | \% Ad- <br> vice change ^^ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |  |  |
| MSY approach $=\mathrm{F}_{\text {MSY }}$ | 1338 | 1258 | 80 | 0.251 | 0.23 | 0.0180 | 5903 | -3.5 | 0.075 | 0.075 |
| Other scenarios |  |  |  |  |  |  |  |  |  |  |
| $F=E U$ MAP ${ }^{\wedge \wedge \wedge}$ : $F_{\text {MSY }}$ | 1338 | 1258 | 80 | 0.251 | 0.23 | 0.0180 | 5903 | -3.5 | 0.075 | 0.075 |
| $F=E U$ MAP $\wedge \wedge \wedge \mathrm{F}_{\text {MSY lower }}$ | 765 | 720 | 45 | 0.136 | 0.126 | 0.0097 | 6504 | 6.3 | -43 | -43 |
| $F=E U$ MAP $\wedge \wedge \wedge \mathrm{F}_{\text {MSY upper }}$ | 2232 | 2097 | 135 | 0.462 | 0.43 | 0.033 | 4963 | -18.9 | 67 | 67 |
| $F=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 7307 | 19.4 | -100 | -100 |
| $\mathrm{F}_{\mathrm{pa}}$ | 1996 | 1877 | 119 | 0.402 | 0.37 | 0.029 | 5212 | -14.8 | 49 | 49 |
| $\mathrm{F}_{\text {lim }}$ | 2531 | 2377 | 154 | 0.543 | 0.50 | 0.039 | 4647 | -24 | 89 | 89 |
| $\mathrm{SSB}_{2024}=\mathrm{Bl}_{\text {lim }}$ | 4919 | 4572 | 347 | 1.59 | 1.48 | 0.114 | 2184 | -64 | 268 | 268 |
| $\mathrm{SSB}_{2024}=\mathrm{B}_{\mathrm{pa}}=\mathrm{MSY} \mathrm{B}_{\text {trigger }}$ | 4063 | 3797 | 266 | 1.10 | 1.02 | 0.078 | 3057 | -50 | 204 | 204 |
| $F=F_{2022}$ | 1312 | 1234 | 78 | 0.25 | 0.23 | 0.0175 | 5930 | -3.1 | -1.87 | -1.87 |
| $\mathrm{SSB}_{2024}=\mathrm{SSB}_{2023}$ | 1130 | 1064 | 66 | 0.21 | 0.193 | 0.0148 | 6120 | 0 | -15.5 | -15.5 |

* Marketable landings, assuming recent discard rate.
** Including BMS landings (EU stocks), assuming recent discard rate.
*** SSB 2024 relative to SSB 2023.
$\wedge$ Total catch in 2023 relative to TAC 2022 ( 1337 tonnes).
^^ Advice value for 2023 relative to the advice value for 2022 ( 1337 tonnes).
$\wedge \wedge \wedge$ EU multiannual plan (MAP) for the Western Waters and adjacent waters (EU, 2019).


Figure 33.1. Sol.27.7fg - Landings and discards estimates by weight, as used by the WG.


Figure 33.2. Sol.27.7fg - InterCatch landings and discard data by year, country and gear.


Age

Figure 33.3a. Sol.27.7fg - Age composition of the catch.


Figure 33.3b. Sol.27.7fg - Age composition of the catch.

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Figure 33.4. Sol.27.7fg - Standardized catch proportion.
















Figure 33.5. Sol.27.7fg - Dotted lines give the length distributions of UK (England and Wales) landings; solid lines of Belgian landings.


Figure 33.6. Sol.27.7fg - Belgian length distributions of discarded and retained fish from discard sampling studies.


Figure 33.7a. Sol.27.7.fg - Mean-standardised indices.


Figure 33.7b. Sol.27.7.fg - Mean-standardised indices.



Figure 33.9. Sol.27.7fg - Mean standardized LPUE UK(E\&W)-BTS-Q3 survey (kg/100 km fished).


Figure 33.10a. Sol.27.7fg - Mean standardised Effort (fishing hours (BE-CBT and IR-CBT), days fished (UK-CBT)).


Figure 33.10b. Sol.27.7fg - Mean standardised LPUE (kg/hour (BE-CBT and IR-CBT), kg/day (UK-CBT)).
biomass indices


Figure 33.11. Sol.27.7fg - Commercial biomass tuning indices.


Figure 33.12. Sol.27.7fg - Summary plots.


Figure 33.13. Sol.27.7fg - One Step Ahead residuals for the final SAM run.


Figure 33.14. Sol.27.7fg - Retrospective analysis.


Figure 33.15. Sol.27.7fg - Comparison with last year's assessment.


Figure 33.16. Sol.27.7fg - Comparison of international TAC, catch and landings.




Figure 33.17. Sol.27.7fg - SAM forecast assuming TAC constraint in the intermediate year followed by targeting $\mathrm{F}_{\text {Msy }}$ in subsequent years.



Figure 33.18. Sol.27.7fg - Year-class sources and contributions for the short-term forecast.


Figure 33.19. Sol.27.7fg - Stock-recruitment plot.

## 34 Sole (Solea solea) in divisions 7.h-k (Celtic Sea South, southwest of Ireland)

## Type of assessment in 2022

No assessment was performed as this is a category 5 stock. No precautionary buffer was applied as it was applied in 2020. ICES advises that when the precautionary approach is applied, catches in 2023 should be no more than 213 tonnes.

## ICES advice applicable to 2021

No assessment was performed as this is a category 5 stock. No precautionary buffer was applied as it was applied last year. Catches in 2022 should be no more than 213 tonnes.
https://www.ices.dk/sites/pub/Publication\ Reports/Advice/2020/2020/sol.27.7h-k.pdf

### 34.1 General

## Stock identity

Sole is a valued, bycatch species in area 27.7 hjk and represents a data-limited stock. This stock was benchmarked for the first time in 2020 as part of WKFlatNSCS (ICES, 2020). During the literature review for this benchmark, no information was found on the identity of this stock. A number of different auxiliary data sources were used to determine the geographical spread and behaviour of this fishery, and where possible its life-history parameters.

Landings data submitted to STECF Fisheries Dependant Information (FDI) (https://stecf.jrc.ec.europa.eu/dd/effort) were used to explore trends in the geographical spread and behaviour of fleets targeting sole in $7 \mathrm{~h}-\mathrm{k}$. Unlike ICES InterCatch data, this data source provides a summary of landings by Member State, gear type, and statistical rectangle. The geographical separation between where the landings in 7 h and 7 j are taken, suggests that there are two discrete fisheries occurring in the stock area (Figure 34.1). This perception is further supported by the clear variation in the gears used to catch sole within the two ICES divisions. Within 7j, sole is predominantly landed by otter trawls, whereas the 7 h fishery is mainly targeted by beam trawls (Figure 34.1 right). This would suggest the two separate assessments are required to effectively manage this fishery.

Due to the data-poor nature of this fishery, there is currently no reliable evidence by which to separate the population of sole in 7 h and 7 j . However, geographical distribution of the landings data would suggest that the fleets are targeting of two discrete populations. Therefore, it is the recommendation of this group to propose sole in $7 \mathrm{~h}-\mathrm{k}$ to the stock identity working group (SIMWG) for further discussion on the possible separation.


Figure 34.1. The spatial distribution of sole landing as reported to the STECF FDI data call in 2016. The landings are plotted by statistical rectangle and show the relative landings reported by Member State (left) and gear type (right), and weighted by the overall landings of sole in ICES divisions 7hjk in 2016.

## Management applicable to 2022

TAC table 2022

| Species: | Common sole <br> Solea solea | Zone:$7 \mathrm{~h}, 7 \mathrm{j}$ and 7 k <br> (SOL/7HJK.) |
| :--- | :---: | :--- | :--- |
| Belgium | 18 | Precautionary TAC <br> Article 8(2) of this Regulation applies |
| France | 36 |  |
| Ireland | 95 |  |
| Netherlands | 28 |  |
| Union | 177 |  |
| United Kingdom | 36 |  |
| TAC |  |  |

## Landings obligation

In 2016, the landings obligation was implemented for this stock for the first time. The regulation (EC, 2015) covered vessels where more than $5 \%$ of their landings using beam trawls were sole during the reference years (2013 and 2014) in ICES divisions 7.b, 7.c and 7.f-7k will be covered by the Landings Obligation. The landings obligation applied to all catches of sole with trammelnets or gillnets. These vessels will have to land all sole in 2016. However, a de minimis exemption will also apply allowing for up to a maximum of $3 \%$ of the annual catch to be discarded. Given the low discards observed in the fishery the landings obligation is unlikely to have a significant impact on this stock or the advice given.

### 34.2 Data

### 34.2.1 Landings and discards

The official and ICES estimates of landings are presented in Table 34.1. The differences between the official and ICES estimated of landing is shown in the unallocated column.

Misreporting is considered to be an issue for this stock but remains difficult to fully quantify. In the past, deviations between official catch statistics and ICES estimates of landings may be due to this misreporting, driven by restrictive TACs in this area ( $7 \mathrm{~h}-\mathrm{k}$ ) and adjoining areas (7e), but also the completeness of data submitted to ICES. Since 2004, these deviations are less due to an improvement in the quality of the data submitted to ICES following on from the introduction of the Data Collection Regulation (EU 2001).

ICES estimates of discards are also provided. Discarding of sole in 7.jk is considered negligible.

### 34.3 Historical stock development

This stock was benchmarked during WKFlatNSCS in 2020 to address the inclusion of available new landings-at-age data for the Division 7.h component, fishery-independent indices, and to consider stock identity (ICES, 2020a). The benchmark concluded that there was no appropriate method for evaluating the stock status and trends, as the sampling only covers a small part of the total fishery, which is not considered representative of the whole area. Therefore, the benchmark agreed to use category 5 to provide advice for this stock.

Prior to the benchmark sole in $\mathrm{h}-\mathrm{k}$ was defined as a category 3 stock as was assessed using an XSA model with commercial landings and LPUE from area 27.7j as data inputs. However, during the benchmark, it was concluded that this stock should be assessed as a category 5 . The age data from the landings-at-age data were presented and indicated that cohort tracking was relatively poor. Landings data were available, but age information was not available from area 27.7 h and precluded using XSA to assess the stock in this area. During previous assessments, it was assumed that the trends in areas 27.7 jk were representative of 27.7 h . The assessment results from the previous assessment were presented during the workshop. The model resulted in relatively poor fits to the data and severe retrospective variability, although the Mohn's rho was within the range of acceptability.
The TAC area includes Division 7h. However, the landings from divisions 7jk are taken in the northeastern part of Division 7 j which is remote from the northern part of Division 7 h , where most of the Division 7h landings are taken. It is likely that the sole from Division 7h are part of the divisions 7e or 7 fg stocks. No further information on stock structure is likely to become available in the short term.

The catches are taken in a mixed fisheries, and should be managed as such. Constraining the landings by TAC will not constrain the catches. Because sole are caught in spatially distinct areas, restricting effort in these areas will be more effective than limiting landings. The catches are taken in a mixed fisheries. and should be managed as such. Constraining the landings by TAC will not constrain the catches. The TAC is currently not restrictive, but for some countries, the quota appears to have become restrictive.

### 34.4 References

ICES. 2020. Benchmark Workshop for Flatfish stocks in the North Sea and Celtic Sea (WKFlatNSCS). ICES Scientific Reports. 2:23. 966 pp. http://doi.org/10.17895/ices.pub. 5976.

EU. 2001. Commission Regulation (EC) No 1639/2001 of 25 July 2001 establishing the minimum and extended Community programmes for the collection of data in the fisheries sector and laying down detailed rules for the application of Council Regulation (EC) No 1543/2000.

Table 34.1. Sole in divisions 7.h-k. History of official landings by country and ICES estimated landings (tonnes).

| Year | Official landings |  |  |  |  |  | ICES landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Belgium | France | UK | Ireland | Other countries | Total |  |
| 1951 | 21 | 150 | 91 | 56 | 0 | 318 |  |
| 1952 | 37 | 220 | 88 | 44 | 0 | 389 |  |
| 1953 | 23 | 227 | 100 | 54 | 0 | 404 |  |
| 1954 | 13 | 317 | 167 | 75 | 0 | 572 |  |
| 1955 | 125 | 634 | 174 | 65 | 0 | 998 |  |
| 1956 | 251 | 511 | 98 | 64 | 0 | 924 |  |
| 1957 | 454 | 359 | 86 | 67 | 0 | 966 |  |
| 1958 | 397 | 605 | 72 | 88 | 0 | 1162 |  |
| 1959 | 241 | 576 | 61 | 101 | 0 | 979 |  |
| 1960 | 0 | 506 | 48 | 96 | 0 | 650 |  |
| 1961 | 197 | 525 | 61 | 110 | 0 | 893 |  |
| 1962 | 144 | 397 | 31 | 123 | 0 | 695 |  |
| 1963 | 149 | 502 | 25 | 127 | 0 | 803 |  |
| 1964 | 310 | 578 | 34 | 118 | 0 | 1040 |  |
| 1965 | 335 | 1128 | 15 | 123 | 0 | 1601 |  |
| 1966 | 123 | 0 | 36 | 118 | 0 | 277 |  |
| 1967 | 168 | 474 | 20 | 123 | 0 | 785 |  |
| 1968 | 113 | 474 | 29 | 116 | 0 | 732 |  |
| 1969 | 175 | 633 | 23 | 120 | 0 | 951 |  |
| 1970 | 436 | 537 | 19 | 122 | 0 | 1114 |  |
| 1971 | 394 | 1382 | 4 | 93 | 0 | 1873 |  |
| 1972 | 203 | 1011 | 11 | 131 | 7 | 1363 |  |
| 1973 | 406 | 390 | 6 | 108 | 4 | 914 |  |
| 1974 | 369 | 143 | 5 | 116 | 15 | 648 |  |
| 1975 | 210 | 207 | 24 | 97 | 2 | 540 |  |


| Year | Official landings |  |  |  |  |  | ICES landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Belgium | France | UK | Ireland | Other countries | Total |  |
| 1976 | 638 | 19 | 11 | 152 | 33 | 853 |  |
| 1977 | 519 | 103 | 12 | 126 | 140 | 900 |  |
| 1978 | 290 | 23 | 11 | 73 | 60 | 457 |  |
| 1979 | 384 | 29 | 18 | 109 | 0 | 540 |  |
| 1980 | 522 | 27 | 42 | 162 | 0 | 753 |  |
| 1981 | 576 | 107 | 83 | 195 | 0 | 961 |  |
| 1982 | 471 | 104 | 108 | 172 | 0 | 855 |  |
| 1983 | 411 | 176 | 129 | 176 | 51 | 943 |  |
| 1984 | 474 | 120 | 151 | 156 | 194 | 1095 |  |
| 1985 | 318 | 25 | 200 | 201 | 280 | 1024 |  |
| 1986 | 442 | 38 | 261 | 188 | 3 | 932 |  |
| 1987 | 271 | 44 | 193 | 168 | 0 | 676 |  |
| 1988 | 254 | 53 | 166 | 182 | 0 | 655 |  |
| 1989 | 252 | 84 | 177 | 206 | 0 | 719 |  |
| 1990 | 353 | 66 | 144 | 266 | 0 | 829 |  |
| 1991 | 358 | 55 | 234 | 306 | 0 | 953 |  |
| 1992 | 312 | 43 | 217 | 255 | 0 | 827 |  |
| 1993 | 317 | 44 | 214 | 237 | 0 | 812 |  |
| 1994 | 338 | 42 | 174 | 184 | 0 | 738 |  |
| 1995 | 433 | 47 | 192 | 243 | 0 | 915 |  |
| 1996 | 375 | 50 | 148 | 182 | 70 | 825 | 443 |
| 1997 | 368 | 58 | 113 | 203 | 0 | 742 | 564 |
| 1998 | 346 | 74 | 111 | 221 | 7 | 759 | 423 |
| 1999 | 101 | 0 | 97 | 207 | 1 | 406 | 381 |
| 2000 | 8 | 78 | 95 | 111 | 10 | 302 | 329 |
| 2001 | 13 | 99 | 111 | 124 | 0 | 347 | 325 |
| 2002 | 154 | 108 | 124 | 129 | 0 | 515 | 430 |
| 2003 | 170 | 133 | 78 | 105 | 0 | 486 | 245 |


| Year | Official landings |  |  |  |  |  | ICES landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Belgium | France | UK | Ireland | Other countries | Total |  |
| 2004 | 157 | 102 | 79 | 111 | 0 | 449 | 454 |
| 2005 | 90 | 93 | 112 | 97 | 0 | 392 | 375 |
| 2006 | 36 | 99 | 88 | 63 | 1 | 288 | 230 |
| 2007 | 31 | 79 | 91 | 77 | 0 | 278 | 232 |
| 2008 | 10 | 58 | 80 | 72 | 0 | 220 | 221 |
| 2009 | 11 | 79 | 58 | 61 | 0 | 208 | 188 |
| 2010 | 20 | 87 | 51 | 71 | 0 | 228 | 206 |
| 2011 | 10 | 95 | 54 | 65 | 0 | 224 | 208 |
| 2012 | 18 | 85 | 46 | 85 | 0 | 234 | 212 |
| 2013 | 4 | 76 | 47 | 85 | 0 | 213 | 204 |
| 2014 | 42 | 61 | 54 | 85 | 0 | 242 | 207 |
| 2015 | 40 | 74 | 53 | 77 | 0 | 244 | 226 |
| 2016 | 91 | 77 | 63 | 99 | 0 | 330 | 269 |
| 2017 | 75 | 81 | 39 | 86 | 0 | 281 | 250 |
| 2018 | 96 | 91 | 33 | 63 | 0 | 283 | 235 |
| 2019 | 75 | 88 | 48 | 55 c | 8 | 274 c | 308 |
| 2020* | 88 | 102 | 50 | 37 | 7 | 284 | 299 |
| 2021* | 50 | 128 | 47 | 75 | 16 | 316 | 336 |

* Preliminary official landings.
c Incomplete due to part of the data being unavailable under national GDPR clauses.


# 35 Whiting (Merlangius merlangus) in Division 6.a (West of Scotland) 

## Type of assessment in 2022

This year's assessment is an update of the procedure used last year (SPALY). Following the decision made earlier during the benchmark meeting WKNSEA 2021 (ICES, 2021a), it was carried out using the state-based assessment model (SAM; Nielsen and Berg, 2014) along with catch and survey data. The assessment followed the procedure outlined in the Stock Annex. Due to COVID-19 disruption and reduced discard sampling from the Nephrops trawl fleet, the estimates of catch numbers for ages 0 and 1 were excluded from the stock assessment. There were no survey data available for the intermediate year as a result of the research vessel breakdown in quarter 1.

A forecast was conducted with short-term stochastic projections according to model and forecast assumptions agreed upon at the WG meeting. These differ to the assumptions agreed at the benchmark and detailed in the Stock Annex as it was believed to be important to account for the change in discarding which appears to have occurred in the fishery since the full implementation of the landing obligation in 2019 (see further details in Section 35.4).

## ICES advice applicable to 2022

"ICES advises that when the MSY approach is applied, catches in 2022 should be no more than 4114 tonnes."
https://www.ices.dk/sites/pub/Publication\ Reports/Advice/2021/2021/whg.27.6a.pdf
Following the benchmark in 2021, the perception of stock status has changed substantially and therefore new advice was issued last year for 2022.

## ICES advice applicable to 2021

"ICES advises that when the precautionary approach is applied, there should be zero catches in each of the years 2021 and 2022."
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2020/2020/whg.27.6a.pdf
Although biennial advice was issued in 2020, the advice for 2022 was revised in 2021 (see "ICES advice applicable to 2022" above).

### 35.1 General

## Stock description

General information is presented in the Stock Annex.

## Management applicable to 2021 and 2022

The TAC for whiting (in tonnes) is set for ICES subareas 6, 12 and 14 and EU and international waters of ICES Division 5b, for 2022-2020 is shown below.

TAC for 2022


| TAC | 1800 |
| :--- | :--- |
| ${ }^{(1)}$ | Exclusively for by-catches of whiting in fisheries for other species. No directed fisheries for whiting <br> are permitted under this quota. |

(Annex IA to Regulation (EU) 2022/109).

## TAC for 2021

The agreed TAC was not available.
TAC for 2020

| Species: | Whiting <br> Merlangius merlangus | Zone: | 6; Union and international waters of 5b; <br> international waters of 12 and 14 <br> (WHG/56-14) |
| :--- | :--- | :--- | :--- |
| Germany | $3\left(^{1}\right)$ | Analytical TAC |  |
| France | 57 | $\left({ }^{1}\right)$ | Article 8 of this Regulation applies |
| Ireland | 273 | $\left({ }^{1}\right)$ | Article 3 of Regulation (EC) No 847/96 shall not apply |
| United Kingdom | 604 | $\left({ }^{1}\right)$ | Article 4 of Regulation (EC) No 847/96 shall not apply |
| Union | 937 | $\left({ }^{1}\right)$ |  |
| TAC | 937 | $\left({ }^{1}\right)$ |  |

$\left({ }^{1}\right)$ Exclusively for by-catches of whiting in fisheries for other species. No directed fisheries for whiting are permitted under this quota.
(Council Regulation (EU) 2020/123).

## Fishery in 2021

A description of the fisheries in the West of Scotland is given in the Stock Annex.
The year 2021 was the third year in a row during which the landing obligation was applied in full force. The overall increase in TAC in 2019-2021 following the introduction of the landing obligation has resulted in an increase in landings.
Total landings (nominal landings, ICES statistics) in 2021 were 851 t, up by $59 \%$ from 2020 (Table 35.1). They were the highest in the last 18 years. The majority were landed by Scottish and Irish vessels, and smaller amounts - by French vessels. The UK landings in Division $6 . a$ in 2021 slightly exceeded the quota for the UK, while Ireland used roughly two-thirds of its quota. Total landings in 2021 constituted $91 \%$ of the TAC for that year.

The total estimated international catch of all age groups in 2021 was 1114 t , of which 261 t were discards (Table 35.2). Of the discards, $73 \%$ were discarded by the trawl fleet targeting crustaceans (Nephrops).

Mandatory introduction of larger square mesh panels for the Nephrops fleet in 2008 seems not to have had much of an effect in the following years on the discards of whiting in Division 6.a. However, in terms of quantity, the discards in 2021 (all ages) were lower (by 69\%) than those in 2020 and they were also below the average in the last decade. In terms of discard rate (discards as a proportion of catch), they were the fifth lowest in the time-series (Table 35.2).

The general perception from fishermen is that large number of whiting are being discarded by the Nephrops fleet and that the numbers of smaller whiting have increased substantially in recent years, but mainly in inshore areas.

### 35.2 Data

## Landings

Total landings, as officially reported to ICES, are shown in Figure 35.1 (in 1965-2021) and Table 35.2 (in 1981-2021).

In the past, there had been concerns that the quality of landings data was deteriorating, giving a possible reason for the different stock dynamics implied by the commercial fleet and the annual survey (ScoGFS-WIBTS-Q1) being in operation at that time (see Section 5.1.6.1.3 in the 2005 WG Report; ICES, 2005). However, a review of previously supplied estimates of misreporting and underreporting (ICES, 2012) carried out at WKNSEA (ICES, 2021a) suggested this to have been a relatively minor issue (in the order of $\sim 5 \%$ of total landings) in the past (since 2001). Therefore, the benchmark agreed that no catch scaling factor for the period 1995-2006 was required in which it differs from previous analytical assessments of this stock.

During WKDEM 2020, the catch data (landings and discards) for 2003 onwards were revised using InterCatch (ICES, 2020a). The age structure in unsampled landings was estimated from that in sampled landings. This was done separately for the two fleets, TR1 (gadoid fishery) and TR2 (Nephrops fishery), on account of the different discard rates observed in them.
The sampling levels in 2021 in the Scottish fleet (taking the majority of the catch), similarly to 2020, were lower compared to the preceding years due to COVID-19. This was particularly the case for discards in the Nephrops fleet. The number of primary sampling units (PSU = number of trips sampled) in the area from 2017 to 2021 is shown in the table below:

| Year | UK (Scotland) |  |  |  | Ireland |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landings |  | Discards* |  | Landings |  | Discards* |  |
|  | TR1 | TR2 | TR1 | TR2 | TR1 | TR2 | TR1 | TR2 |
| 2021 | 11 | 0 | 8 | 4 | 80 | 0 | 32 | 0 |
| 2020 | 11 | 1 | 12 | 4 | 28 | 0 | 20 | 0 |
| 2019 | 16 | 0 | 18 | 23 | 23 | 0 | 28 | 0 |
| 2018 | 14 | 1 | 11 | 30 | 28 | 0 | 60 | 0 |
| 2017 | 11 | 0 | 13 | 39 | 23 | 0 | 48 | 0 |

[^21]Landings uploaded to InterCatch by métier and country for 2021 are shown in Figure 35.2. As in previous assessments, age distributions were estimated from market samples. Total catch (including landings) by métier for 2021 is shown in Figure 35.3. Catch numbers-at-age (in different catch categories) in 2021 are shown in Figure 35.4. Last year, no landings were recorded for fish at age 1 and this was a consequence of the absence of this age group in the sampled landings. The highest numbers in landings were found for age groups $2-5$. Catch numbers-at-age from 2003 onwards are shown in Figure 35.5.

Annual numbers-at-age in the landings are given in Table 35.3. Annual mean weights-at-age in the landings are given in Table 35.6 and shown in Figure 35.6. Last year, they decreased slightly in most age groups. Overall, the mean weights-at-age in the landings have been variable in recent years due to the variability associated with low sample sizes. Efforts to increase sampling in these fisheries are being pursued.

## Discards

This WG's estimates of discards have been based on data collected in the Irish and Scottish discard programme and raised by landings. Discard age compositions from Scottish and Irish samples have been applied to unsampled fleets. As agreed at WKDEM 2020, the raising and age allocations for discards were done separately for the TR1 and TR2 fleets (ICES, 2020a).
Discards uploaded to InterCatch by métier and country for 2021 are shown in Figure 35.2.
Annual numbers-at-age in the discards are given in Table 35.4. Due to the low sample numbers in the Nephrops fleet, the resulting numbers at age 0 and 1 in discards were considered as unreliable (disproportionately low), especially when compared with previous years (Figure 35.7). This had further consequences for the assessment (see section '35.3 Stock assessment' below).

Annual mean weights-at-age in the discards are given in Table 35.7 and shown in Figure 35.6. Mean weights-at-age 1 in discards were considerable higher compared to previous years, but this was most likely a result of the low number of samples from the Nephrops fleet.

## Biological

Annual numbers-at-age in the total catch are given in Table 35.5. Annual mean weights-at-age in the total catch are given in Table 35.8 and shown in Figure 35.6.

In previous assessments prior to 2021, mean catch weights-at-age were used as mean stock weights-at-age. Since 2021, the latter has been estimated using the method elaborated at WKNSEA 2021 (ICES, 2021a) that combines catch and survey weights-at-age (see the Stock Annex). Two sets of stock mean weights-at-age are delivered: one to be used as stock weights-atage input into the SAM stock assessment model to calculate SSB, and one to be used to estimate size-dependent natural mortality-at-age. The estimates from the former (smoothed with a General Additive Model, GAM) are shown in Table 35.9 and Figures 35.6 and 35.8.

In previous assessments of whiting in Division 6.a, natural mortality was assumed to vary and be dependent on fish weight (Lorenzen, 1996). $M$ values were time-invariant and were calculated as:

$$
M_{a}=3.0 \bar{W}_{a}^{-0.29}
$$

where $M_{a}$ is natural mortality-at-age $a, \bar{W}_{a}$ is the mean stock weight-at-age $a$ (in g ) and the numbers are the Lorenzen's parameters for fish in natural ecosystems.

During WKNSEA in 2021 it was agreed to first smooth the time-series of stock mean weights-atage using a GAM and then use these smoothed weights-at-age in the Lorenzen (1996) equation to obtain a time-series of mortality-at-age estimates to be used as input in the stock assessment model (ICES, 2021a).

The time-series of smoothed stock mean weights-at-age obtained from a combination of catch data and survey data from Q1 and Q4 are used in the Lorenzen equation. The smoothed stock mean weights-at-age are shown in Figure 35.6. These estimated natural mortality-at-age is shown in Table 35.10 and Figure 35.9.

In earlier assessments prior to 2021, maturity-at-age was assumed to be knife-edge with the value 0 at age 1 and full maturity-at-age $2+$. An analysis of Scottish survey data conducted at WKDEM 2020 and updated at WKNSEA 2021, showed no clear temporal trends in maturity (ICES, 2020a; ICES, 2021a). The analysis provided coefficients of the logistic model (being time-invariant, with data up to 2020): -6.307 (intercept) and 5.228 (slope). The midpoint of the modelled maturity ogive, A50, was estimated to be $1.206( \pm 0.031)$ years. The estimated proportions of mature whiting are shown in the table below:

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | 7+ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Maturity ogive | 0 | 0.254 | 0.984 | 1 | 1 | 1 | 1 | 1 |

The analysis revealed that considerable proportion of fish at age 1 (a quarter) and nearly all fish at age 2 were mature. There was little variability in the data resulting in relatively narrow confidence intervals (ICES, 2021a).

## Surveys

Five research vessel survey series for whiting in 6 .a were available to the WG in previous years. They included the two 'old' Scottish surveys:

- Scottish first-quarter west coast groundfish survey (ScoGFS-WIBTS-Q1): all ages 1 and older, years 1985-2010;
- Scottish fourth-quarter west coast groundfish survey (ScoGFS-WIBTS-Q4): all ages including age 0, years 1996-2009.

The Q1 Scottish Groundfish Survey was performed using a repeat station format with the GOV survey trawl together with the west coast groundgear rig ' $C^{\prime}$. The Q4 Scottish Groundfish Survey also used the GOV survey trawl with groundgear ' $\mathrm{C}^{\prime}$ and the fixed station format. The Q 4 survey was not carried out in 2010 due to an engine breakdown of the research vessel.

In 2011, the Q1 and Q4 Scottish Groundfish Surveys were re-designed. The previous repeat station survey format consisting of the same series of survey trawl positions being sampled at approximately the same temporal period every year is considered a rather imprecise method for surveying both these subareas. Therefore, a move towards some sort of random stratified survey design was judged necessary (see further details of the modified survey design in the Stock Annex). The introduction of the new design initiated two 'new' time-series:

- Scottish first-quarter west coast groundfish survey (UK-SCOWCGFS-Q1): all ages 1 and older, years 2011-2021;
- Scottish fourth-quarter west coast groundfish survey (UK-SCOWCGFS-Q4): all ages including age 0, years 2011-2021.

The distribution and densities of whiting at-age (standardised as CPUE) in the Q1 and Q4 surveys in 2017-2021 are shown in Figure 35.10. The Q4 survey in 2013 was not fully implemented due to adverse weather conditions. It covered only the northern half of Division 6.a and therefore, the index for that year was not used in assessments prior to 2020. Due to vessel breakdown, the Q1 survey was not carried out in 2022. As a result, 11 years of data are currently available in the time-series for the Q1 survey and ten years of data for the Q4 survey (as valid indices).
The Irish Groundfish Survey has partly been conducted in Division 6.a:

- Irish fourth-quarter west coast groundfish survey (IGFS-WIBTS-Q4): all ages including age 0, years 2003-2021.

The distribution and densities of whiting at-age in the two Q4 surveys, UK-SCOWCGFS-Q4 and IGFS-WIBTS-Q4 in 2017-2021 are shown in Figure 35.11 (only the southern part of Division 6.a). The Irish survey uses the RV Celtic Explorer and is part of the IBTS coordinated western waters surveys. The vessel uses a GOV trawl, and the design is a depth-stratified survey with randomised stations. Effort is recorded in terms of minutes towed. The previous Irish survey (IreGFS), being in operation in 1993-2002 (see the Stock Annex), is not used anymore in the assessment.
Further descriptions of the above five surveys can be found in 'Manual of the IBTS North Eastern Atlantic Surveys' (ICES, 2017) and in the last IBTSWG report (ICES, 2021b).
During WKNSEA 2021, it was agreed to combine all the three Q4 surveys (ScoGFS-WIBTS-Q4, UK-SCOWCGFS-Q4 and IGFS-WIBTS-Q4) into one survey index for use in the stock assessment (ICES, 2021a). The analysis of the combined index was conducted using a GAM-based deltalognormal model Berg et al. (2014) including a number of explanatory variables. The combined index (denoted as Comb-WCGFS-Q4) derived from the model fit is shown in Figure 35.12. The index provides a more complete representation of the population compared to the respective indices used on their own. It simplifies the modelling procedure in the annual assessments of the stock (with three rather than five indices) and provides a longer continuous time-series.

## Commercial cpue

Four commercial catch-effort time-series were previously available to the WG, but they have not been used for a number of years. They are only presented in the Stock Annex.

### 35.3 Stock assessment

In the years 2011-2019, the assessment was done using a Time-Series Analysis (TSA) model (Gudmundson, 1994; Fryer, 2002; Needle and Fryer, 2002). At that time, the stock was classified as category 1. During the benchmark process of WKDEM 2020, it was found that running TSA with the new data and changed survey configuration posed a challenge (ICES, 2020a). Poorly converged optimisation runs (with some parameters being found on the boundary of the assumed parameter space) in conjunction with excessive running times were a major obstacle to complete the assessment successfully. In these circumstances, it was decided ad hoc to run the benchmark assessment using an alternative method; namely, a SPiCT model (Pedersen and Berg, 2017). At the same time, the stock was downgraded to category 3 and further to category 5 according to the ICES guidelines for data-limited stocks (ICES, 2019).

In the benchmark process of WKNSEA 2021, it was decided to use SAM as the assessment method (ICES, 2021a). It was agreed that the model should be run over the entire time period for which catch numbers-at-age data were available in order to capture the earliest part of the timeseries (during which catches were relatively high). To facilitate this in SAM, it was assumed that catch and discards mean weights-at-age zero between 1981 and 2002, and landings mean
weights-at-age zero for the entire modelled time period, were equal to the average of mean weights-at-age zero between 2003 and 2020. In addition, stock mean weights-at-age and natural mortality-at-age between 1981 and 1984 were assumed to equal estimates for the equivalent quantity from the earliest available year (i.e. 1985). Catch numbers-at age zero are only available from 2003 onwards (from the WKDEM data call) and therefore values between 1981 and 2002 were treated as missing and estimated in the assessment model.

## Data screening

The diagnostics for commercial catch data and the three indices considered as tuning series (ScoGFS-WIBTS-Q4, UK-SCOWCGFS-Q1 and Comb-WCGFS-Q4) for the assessment are shown in Figures 35.13-35.15).

The log catch curves for the commercial catch and for the surveys in the current assessment are shown in Figure 35.13. In most cases, the curves are relatively linear and not very noisy. They also show a fairly steep and consistent drop in abundance. The curves for the commercial catch have a strong 'hook', especially at age 0 (from 2003 onwards).

The plots of mean standardised catch proportions at age by year (Figure 35.14) demonstrate that there is some general consistency in the estimates of year-class strength across age groups. They indicate strong year classes in recent years (2009 and 2014 year classes), but also markedly weak year classes (2012 and 2017 year classes). A clear year effect can be seen for ages $1+$ in 2007 and 2008.

The within-survey correlation plots generally show significant correlations between consecutive age groups (Figure 35.15). There is a general consistency in the estimates of year-class strength across age groups, but the points are more scattered for old age groups.
The three indices used as tuning series in the current assessment are shown in Table 35.11.

## Final assessment

Model used: SAM
Software used: stockassessment package in R; stockassessment.org
Input data types and characteristics:

- Catch numbers-at-age: ages 2-7+, 1981-2021; age 1, 1981-2020; age 0, 2003-2019
- Landings fraction at age, ages 0-7+, years 1981-2021,
- Catch weights-at-age: ages 1-7+, 1981-2021; age 0, 2003-2021 (excl. 2020, see below)
- Landings weights-at-age, ages 0-7+, years 1981-2021 (excl. age 1 in 2021, see below),
- Discards weights-at-age: ages 1-7+, years 1981-2021; age 0, 2003-2021 (excl. 2020, see below)
- ScoGFS-WIBTS-Q1, ages 1-6, years 1985-2010,
- UK-SCOWCGFS-Q1, ages 1-6, years 2011-2021,
- Modelled Q4 index, Comb-WCGFS-Q4, fitted to data from ScoGFS-WIBTS-Q4 (G4299), UK-SCOWCGFS-Q4 (G4815), and IGFS-WIBTS-Q4 (G7212); ages 0-7+ with variance estimates; 1996-2021.

No age-0 discards were recorded in 2020. This is assumed to be due to a lack of discard sampling in the Nephrops fishery from Q2 onwards and therefore age zero discards are treated as missing in 2020 (as in the years before 2003). Missing catch weights-at-age zero are assumed equal to the average of 2003-2019. There were no age-one fish recorded in the landings in 2021, therefore landings weight-at-age one for 2021 is a three-year average value (2018-2020). Catch numbers-
at-age zero and one for 2021 were considered unreliable due to low sampling levels, and were thus removed from the dataset. Unknown/removed catch numbers-at-age values are estimated in the model. Typically, survey indices for UK-SCOWCGFS-Q1 would be available for the assessment year. The index was not calculated in 2022 as there was no survey carried out due to vessel breakdown. The 2022 assessment thus runs up to 2021, deviating from the model configuration agreed in the benchmark.

The assessment of whiting in $6 . a$ was conducted using a SAM model fitted to the updated catch and survey dataseries. Full details of the model implementation are presented in the Stock Annex. The SAM configuration file for the final assessment model run is given in Table 35.12. To summarise the main configuration settings:

- Fishing mortality states processes are uncoupled across all age groups.
- Catchabilities for each survey index are freely estimated with the exception of the two oldest age groups for each index; ages 5 and 6 in ScoGFS-WIBTS-Q1 and UK-SCOWCGFS-Q1, and ages 6 and 7+ in Comb-WCGFS-Q4.
- Catch observation variance parameters are allowed to differ for age 0 and age $7+$ while all other age groups are coupled.
- Survey observation variance parameters are coupled across all ages for ScoGFS-WIBTSQ1 and UK-SCOWCGFS-Q1, whereas for Comb-WCGFS-Q4 observation variance parameters were uncoupled for age zero, and coupled for ages 1 to 4 and ages 5 to 7+.
- The catch, ScoGFS-WIBTS-Q1, and UK-SCOWCGFS-Q1 fleets are modelled with independent covariance structures, whereas the Comb-WCGFS-Q4 fleet is modelled with a first order autoregressive variance structure (AR1) with ages 0 and 1, ages 1 to 6 , and ages 6 and 7+ coupled.
- Recruitment is modelled as a random walk.
- $\bar{F}$ was calculated for ages 1 to 3 in order to reflect changes in fishery selectivity, moving from a target fishery in the 1980s and 1990s to a bycatch and discard component of the Nephrops trawl fishery from the early 2000s onwards. This is a change from previously accepted analytical assessments of this stock which used an $\bar{F}$ range of ages 2 to 4 .

Table 35.13 shows the SAM parameter estimates for the assessment model. Table 35.14 shows the population numbers-at-age estimated in SAM, and estimated F-at-age is shown in Table 35.15. A summary of the full model output is detailed in Table 35.16. The summary plots for the final assessment are shown in Figure 35.16.

The fits of the model to observations (catch and survey indices on a log scale) are shown in Figure 35.17. The fits to the ScoGFS-WIBTS-Q1 appear better at younger ages while for ages five and six some of the trend in the early part of the time-series is not captured. Fits to Comb-WCGFS-Q4 are also generally good, although there is a tendency towards overestimation of age zero individuals for part of the time-series, and some deficiency in tracking the variability of age seven+. The shorter time-series of UK-SCOWCGFS-Q1 makes it more difficult to assess the model fit in terms of trends, but the model seems to fit the observations reasonably well. The model also appears to follow the catch data well for most ages, but perhaps tracks less of the interannual variability for ages 6 and 7+.

The diagnostics of the quality of the model fit were: examination of the residuals; a leave-oneout analysis of the relative influence of indices on model estimates; a retrospective peel analysis. One observation ahead residuals-at-age for catch and survey indices are shown in Figure 35.18. The residuals were not substantially affected by the updates made to the model, showing similar patterns to last year's assessment model. There is an observable trend in the catch residuals from the late 1990s to mid-2000s, particularly between ages two and four, where the fishery shifted from being directly targeted to bycatch. There is some tendency towards negative residuals in
the oldest age class of the Q4 survey index, but they are still occasionally interspersed with positive residuals. Otherwise, there are no particularly problematic trends in magnitude or direction.

The model leave-one-out analysis is shown in Figure 35.19. Exclusion of each index in turn results in estimates of SSB, $\bar{F}$ and recruitment which follow very similar trends over time, suggesting generally good agreement between indices. Estimates of SSB in more recent years have a tendency towards being generally lower with the exclusion of ScoGFS-WIBTS-Q1 and Comb-WCGFS-Q4, while remaining reasonable stable with the exclusion of UK-SCOWCGFS-Q1, when compared to the final model. Leave-one-out estimates remain within the confidence interval of full model estimates in all cases. Estimates of $\bar{F}$ in each case diverge the most between 2000 and 2010. Excluding the old Q1 survey (ScoGFS-WIBTS-Q1) results in higher estimates of $\bar{F}$ for much of the time-series from 2000 onwards, while excluding the Q4 index (Comb-WCGFS-Q4) results in generally lower estimates of $\bar{F}$ for the same period.

Retrospective peels for the updated assessment model are shown in Figure 35.20. Retrospective bias in SSB is not substantial, with some downward revision with the addition of new data in recent years. The Mohn's rho values are as follows:

| SSB | $\bar{F}$ | Recruitment |
| :---: | :---: | :---: |
| 0.17 | -0.14 | 0.41 |

The relatively high Mohn's rho value for recruitment is a result of the consistently low recruitment values estimated in recent years. Only one recruitment peel falls outside the confidence interval envelope.

A sensitivity analysis was also carried out to assess the effect of omission of catch-at-age zero and one data from the model. Retrospective fits for the past five years were estimated based on censored datasets, with catch-at-age zero and one values being freely estimated in SAM. These estimates were compared to the retrospective fits for the full model (Figure 35.21). Model estimates appear robust to removal of catch numbers-at-age data for that period, with the only notable divergence being where SAM estimates a lower $\bar{F}$ in 2021 (dotted black line) based on the unreliable catch numbers-at-age zero and one data, when compared to the equivalent value of $\bar{F}$ which was freely estimated in the data censored model (solid black line).

The SAM stock-recruit plot is presented in Figure 35.22 and suggests a relationship which has experienced a number of reasonably distinct phases over time. SSB and recruitments were relatively high, but decreasing, in the early 1980s. At the latest benchmark, it was suggested that this phase was related to the gadoid outburst of the 1960s and 1970s, and the decreasing stock size at the beginning of the modelled period is the time at which the population was returning to its usual size (Holden, 1991; Hislop, 1996). Stock size was then relatively stable for much of the 1990s, but declined in the early 2000s. SSB has shown an increasing trend since $\sim 2010$ and average recruitment since then is higher than in the previous ten years.

## Comparison with last year's assessment

The 2021 assessment was carried out using SAM, following a benchmark of the stock. The 2021 assessment model was revised in 2022 due to an error in stock mean weights-at-age data inputs which affected SSB estimates, but not the published catch advice.

The revised estimates for last year's assessment are as follows:
$\bar{F}_{(1-3)}$ in $2020=0.065$,

SSB in $2021=30357 \mathrm{t}$.
The estimated fishing pressure continued to be very low. The stock biomass was estimated to increase slightly in 2021.

In this year's assessment estimates were updated and remain very similar to last year with some slight revisions (Figure 35.23).

## SURBAR analysis

An alternative exploratory assessment conducted using SURBA (Needle, 2015) was presented at the WKNSEA benchmark (ICES, 2021a; WD 5.5 Whiting 6a SURBAR) and WGCSE 2021. Its updated run was presented to the WG this year.

This method requires stock weights-at-age, maturity ogive and survey indices. The smoothed estimates of stock weights-at-age were deployed in the model (those used to calculate SSB and shown in Table 35.9). The same three tuning series were considered for the model for SAM:

- ScoGFS-WIBTS-Q1 for the period 1985-2010;
- UK-SCOWCGFS-Q1 for the period 2011-2021;
- Comb-WCGFS-Q4 for the period 1996-2021.

The model used the following settings:

- Three survey series (as above);
- $\quad$ Reference age for separable model = 3;
- Lambda smoother = 1.0;
- $\quad$ All SSQ weightings and catchabilities q set to 1.0.

The model produced the output given in Figures 35.24 . The stock summary plots show rather variable estimates of mean $Z$ being generally lower from the mid-2000s onwards. SSB rose to a peak in the mid-1990s, before returning back down to the levels seen in the late 1980s with a substantial increase in the recent period. Also, it seems to fluctuate more in recent years compared to the historical period. The increase between 2019 and 2021 can be explained by relatively high recruitments in 2018-2019 and very low mean Z (associated with almost flat catch curves between 2019 and 2021 across a number of cohorts in UK-SCOWCGFS-Q1, Figure 35.13). Recruitment (at age 1) between 2005 and 2013 remained on a very low level. In recent years, it has increased markedly.

The assessment with SURBAR shows similar trends to those seen in the SAM runs (Figure 35.25).

## State of the stock

The spawning-stock biomass (SSB) in 2022 is estimated to be above MSY $B_{\text {trigger. }}$. It has increased since 2010 and is now at a level consistent with that estimated for the late 1990s. (Figure 35.26). $\bar{F}$ declined almost continuously since from 2000-2008, and has been below Fmsy since 2005. Recruitment was at historically low levels in the mid-2000s, and since around 2014 has been variable around relatively low values. The recruitment in 2021 was estimated to be at the median level of the past decade.

### 35.4 Short-term projections

The WG conducted a forecast using SAM in the form of short-term stochastic projections. A total of $1 \times 10^{5}$ samples was generated from the estimated distribution of survivors. These replicates were then simulated forward according to model and forecast assumptions (see below), using
the usual exponential decay equations, but also incorporating the stochastic survival process (using the estimated survival standard deviation) and subject to different catch-options scenarios.

Recruitment in the intermediate year (2022) would typically be assumed to equal the SAM estimate, however, due to a lack of Q1 survey data in 2022, it was assumed to equal the median resampled value from 2012-2021, reflecting recent levels of recruitment. The estimate of recruitment for the forecast year was also assumed to equal the median resampled value from 20122021 as per the stock annex.

Fishing mortality in the intermediate year (2022) was taken as a five-year average over 2017 to 2021 (Figure 35.27) of the exploitation pattern rescaled to the 2021 mean F.

The stock has been subject to the landings obligation since 2019, at which point a bycatch TAC of 1112 t was set to allow fisheries with a whiting bycatch component to continue (this represented an increase from the 213 t TAC set for the preceding three years). This increased TAC appears to have resulted in a change in discarding practices since 2019. In Figure 35.28, the observed proportion discarded at-age shows a significant decline in the proportion of discards for ages 3 to 5 between 2019 and 2021, as well as a decline in proportion of discards at age 2 over the same period. For the forecast, total catch is partitioned into landings and discards on the basis of the mean discard proportions-at-age between 2019 and 2021 (rather than the more five-year average agreed at the benchmark and documented in the Stock Annex) with the assumption that this observed change in behaviour will continue in 2022 and 2023.

The observed mean weights-at-age zero and one in the 2021 discard data were considered unreliable due to limited sampling of fleets discarding whiting (COVID-19 pandemic related sampling issue, see Section 35.2 for further details). For the purpose of forecasting, the mean value for those ages should be taken from 2019-2020 data only.

| Variable | Value | Notes |
| :--- | :--- | :--- |
| $\mathrm{F}_{\text {ages 1-3 (2022) }}$ | 0.07 | $\mathrm{~F}=\mathrm{F}_{\text {average (2017-2021) }}$ rescaled to $\mathrm{F}_{2021}$ |
| SSB (2023) | 28727 | Short-term forecast; Tonnes |
| $\mathrm{R}_{\text {age } 0 \text { (2022-2023) }}$ | 273676 | Median recruitment, resampled from the years 2012-2021; thousands |
| Catch (2022) | 1484 | Short-term forecast using an F F 2021 ; tonnes |
| Projected landings (2022) | 766 | Short-term forecast; assuming average landings ratio by age 2019- <br> 2021; tonnes |
| Projected discards (2022) | 718 | Short-term forecast; assuming average discard ratio by age 2019-2021; <br> tonnes |

Under the forecast assumption of status quo F, landings in 2022 are predicted to be 766 t and discards to be 718 t . The SSB in 2023 is forecast to be 28727 t , which is above $\mathrm{B}_{\lim }$ and MSY $\mathrm{B}_{\text {trigger }}$. A summary of the forecast run under different catch scenarios for 2023 is shown in Table 35.17 (the values that appear in the catch scenarios are medians from the distributions that result from the stochastic forecast).

The forecast stock trajectory under the proposed advice for 2023 ( 4155 t ) shows a decrease in SSB in 2024 (Figure 35.29). Figure 35.30 shows the contribution by recruitment year to SSB in 2024 and catch in 2023 (when fished at $\mathrm{F}_{\text {MSY }}$ ). The assumption regarding recruitment in 2022 and 2023 contribute approximately $26 \%$ and $9 \%$ to the forecast 2024 SSB, and $49 \%$ and $6 \%$ to the forecast 2023 catch, respectively.

### 35.5 MSY and biological reference points

The reference points for this stock were updated at WKNSEA 2021 (ICES, 2021a), following the general approach agreed at WKMSYREF4 (ICES, 2016).

The reference points estimated in 2021 are summarised in the table below:

| Reference point | WKMSY-REF4 2016 | $\begin{aligned} & \text { WGCSE } \\ & 2016 \end{aligned}$ | $\begin{aligned} & \text { WKNSEA } \\ & 2021 \end{aligned}$ | Rationale (WKNSEA 2021); details |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{Bl}_{\text {lim }}$ | 28500 t | 31900 t | 17286 | Lowest SSB (1999) within period of high recruitment (pre 2000) |
| $\mathrm{B}_{\mathrm{pa}}$ | 39900 t | 44600 t | 25597 | Blim $x \exp (1.645 \times \sigma) ; \sigma=0.239$ (CV on estimate of SSB2020) |
| $\mathrm{F}_{\text {lim }}$ | 0.25 | 0.27 | 0.31 | F giving 50\% probability of SSB<Blim in stochastic simulation (EqSim) Uses segmented regression recruitment with breakpoint=Blim (S-R pairs from 1985 onwards). |
| $\mathrm{F}_{\mathrm{pa}}$ | 0.18 | 0.19 | 0.21 | Fp.05; the F that leads to SSB $\geq$ Blim with 95\% probability. |
| $\mathrm{F}_{\text {MSY }}$ | 0.22 | 0.23 | 0.21 | Fp. 05 (FMSY uncapped $=0.23$ ) |
| $\mathrm{F}_{\text {MSY }}$ upper | 0.34 | 0.32 | 0.21 | Fp. 05 (FMSY upper uncapped $=0.27$ ) |
| $\mathrm{F}_{\text {MSY }}$ lower | 0.16 | 0.15 | 0.173 | F resulting in no more than 5\% reduction in longterm yield compared with MSY without ICES AR (95 \% yield at Fp.05). |

### 35.6 Management plans

There are no specific management objectives or a management plan for this stock, but the EU multiannual plan takes bycatch of this species into account (EU, 2019).

### 35.7 Uncertainties and bias in the assessment and forecast

Some uncertainties signalled in previous assessments were related to area misreporting of landings. Marine Scotland Compliance have provided estimates based on their surveillance and monitoring programme which suggest area misreporting of whiting to be in the order of $10-15 \%$ of reported landings in recent years (ICES, 2012). This issue is thus considered to be of relatively minor importance.

As a result of the 2021 benchmark, the stock was changed from a category 5 stock to a category 1 stock. The assessment, which is now based on SAM, includes revised catch and survey data, updated biological parameters, and accounts for changes in fishery selectivity (rather than changes in survey catchability). These changes have resulted in an improved assessment of the stock status.

The retrospective bias observed for SSB and recruitment is potentially an issue; the Mohn's rho has been found outside the bounds suggested by WKFORBIAS (ICES, 2020b). However, the assessment is deemed to be valid and provide advice based on the WKFORBIAS decision tree.

The lack of an intermediate year survey in Q1 2022 required a change to the recruitment assumptions and this increased the uncertainty of the forecast.

The sampling levels both in 2020 and in 2021 were lower compared to previous years due to COVID-19. As a result, total discards were underestimated for 2021, and the estimates of catch
numbers for ages 0 and 1 were excluded from the stock assessment. Sensitivity analyses indicate that these issues are likely to have minimal impact on the assessment.

### 35.8 Recommendation for next benchmark

Although the combined Q4 index is considered as representative of the population, there is scope for its further improvement. During WKNSEA 2021, the potential need for inclusion of an interaction term between year and geographical coordinates was discussed. Exploratory analysis suggested some temporal changes in the distribution of age groups, but it was found that the inclusion of the interaction term had little effect on the index values or the internal consistency. Additional analyses and careful sense checking of estimated covariates would be necessary to find optimal settings for such an augmented model.
The Q1 indices used in the assessment represent CPUE calculated for age groups. While they integrate information for specific areas or strata, other sources information are thereby ignored. It seems plausible that these indices can be improved by including other explanatory variables, in a similar way as for the Q4 index.

Alternative approaches could be used to estimate stock mean weights. Currently, they are smoothed with a GAM independently in each age group. As a result, two consecutive age groups can occasionally show different trends, which was observed here and which makes such estimates more uncertain.

While a Random Walk on recruitment was selected for the assessment for practical reasons, alternative approaches should be explored in modelling the S-R relationship model with internal calculation of reference points.

### 35.9 Management considerations

SSB in 2022 is estimated to be above MSY Btrigger. It has increased since 2010 and is now at a level consistent with that estimated for the late 1990s. Fishing pressure (F) declined almost continuously since around 2000 and is below Fmsy since 2005. After a period of somewhat higher recruitment (2014-2019), recruitment in the two last years was estimated to be lower. The first non-zero advice has been given for five years. The SSB is forecast to decline in the following two years in all catch scenarios.

Whiting are caught and heavily discarded in small-meshed fisheries for Nephrops. Under the landing obligation (since 2019), discards have considerably reduced. However, they still make up a considerable proportion of the catch ( $23 \%$ in 2021). Reported BMS landings are negligible; they are much lower than ICES estimates of catches below minimum size (i.e. estimated discards at age $0-2$ ).

TAC increased under the landing obligation to allow continuation of mixed fisheries. In response to this, the discard rate declined. The forecast assumes that this behaviour will continue in 2022 and 2023.

It should be noted that TAC have been set for a larger area than Division 6.a. and include areas 6.b, 5.b and international waters of 12 and 14 (Annex IA to Regulation (EU) 2022/109).

Whiting are caught in mixed fisheries with cod and haddock in Division 6.a. There have been several technical conservation measures introduced in the 6.a gadoid fishery in recent years. The increase in mesh size from 100 mm to 120 mm , established under the emergency measures since 2010, and the introduction of large square mesh panels in the Nephrops fishery, are likely to have contributed to the observed reductions in fishing mortality.

### 35.10 References

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Table 35.1. Whiting in Division 6.a. Nominal landings (in tonnes) as officially reported to ICES.

| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 1 | - | + | - | + | + | + | - | 1 | 1 | + | - | - | - | - | + | - | - | - | - |
| Denmark | 1 | + | 3 | 1 | 1 | + | + | + | + | - | - | - | - | - | + | + | - | - | - | - |
| Faroe Islands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | + | + |
| France | 199 | 180 | 352 | 105 | 149 | 191 | 362 | 202 | 108 | 82 | 300 | 48 | 52 | 21 | 11 | 6 | 9 | 7 | 6 | 1 |
| Germany | + | + | + | 1 | 1 | + | - | + | - | - | + | - | - | - | - | - | - | + | 1 | - |
| Ireland | 1315 | 977 | 1200 | 1377 | 1192 | 1213 | 1448 | 1182 | 977 | 952 | 1121 | 793 | 764 | 577 | 568 | 356 | 172 | 196 | 56 | 69 |
| Netherlands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Norway | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Spain | - | - | - | - | - | - | 1 | - | 1 | 2 | + | - | 2 | - | - | - | - | - | - | - |
| UK (E, W and NI) | 44 | 50 | 218 | 196 | 184 | 233 | 204 | 237 | 453 | 251 | 210 | 104 | 71 | 73 | 35 | 13 | 5 | 2 | 1 |  |
| UK (Scot.) | 6109 | 4819 | 5135 | 4330 | 5224 | 4149 | 4263 | 5021 | 4638 | 3369 | 3046 | 2258 | 1654 | 1064 | 751 | 444 | 103 | 178 | 424 |  |
| UK (total) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 370 |
| Total landings | 7669 | 6026 | 6908 | 6010 | 6751 | 5786 | 6278 | 6642 | 6178 | 4657 | 4677 | 3203 | 2543 | 1735 | 1365 | 819 | 289 | 383 | 488 | 441 |

## Table 35.1. Continued.

| Country | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020* | 2021* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Denmark | - | - | - | - | - | - | - | - | - | 2 | 56 | 10 | - |
| Faroe Islands | - | + | 1 | 1 | - | - | - | - | - | - | - | - | - |
| France | 1 | 3 | + | + | 1 | 1 | + | 5 | 3 | 2 | 7 | 10 | 35 |
| Germany | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Ireland | 125 | 99 | 149 | 96 | 97 | 97 | 88 | 77 | 53 | 72 | ** | 126 | 161 |
| Netherlands | - | - | - | - | - | - | 11 | 52 | 19 | 4 | 23 | 4 | + |
| Norway | 2 | - | - | - | - | - | - | - | - | + | - | - | - |
| Spain | - | - | - | - | - | - | - | - | - | - | + | - | + |
| UK (E, W and NI) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| UK (Scot.) |  |  |  |  |  |  |  |  |  |  |  |  | - |
| UK (total) | 354 | 247 | 80 | 204 | 116 | 83 | 122 | 98 | 94 | 108 | 241 | 387 | 654 |
| Total landings | 482 | 349 | 230 | 301 | 214 | 181 | 221 | 232 | 169 | 189 | $327^{* *}$ | 537 | 851 |

## * Preliminary.

** Incomplete/missing due to part of the data being unavailable under national GDPR clauses.
$+<0.5 \mathrm{t}$.

Table 35.2. Whiting in Division 6.a. Landings, discards and catch estimates for 1981-2020, as used by the WG. Values are totals for ages 1 to 7+ (in 1981-2002) and for ages 0 to 7+ (in 2003-2020). Discard and catch values for the years 19812002 are revised compared to previous assessments because of a revised method for raising discards (Millar and Fryer, 2005). Landings, discard and catch values for the years 2003-2018 are revised InterCatch estimates (ICES, 2020a) compared to previous assessments.

| Year | Landings | Discards | Total | Discard rate (\%) |
| :---: | :---: | :---: | :---: | :---: |
| 1981 | 12203 | 2132 | 14335 | 15 |
| 1982 | 13871 | 5485 | 19356 | 28 |
| 1983 | 15970 | 6294 | 22264 | 28 |
| 1984 | 16458 | 4017 | 20475 | 20 |
| 1985 | 12893 | 4840 | 17733 | 27 |
| 1986 | 8454 | 2669 | 11123 | 24 |
| 1987 | 11544 | 11918 | 23462 | 51 |
| 1988 | 11352 | 8132 | 19484 | 42 |
| 1989 | 7531 | 5876 | 13407 | 44 |
| 1990 | 5643 | 4530 | 10173 | 45 |
| 1991 | 6660 | 4883 | 11543 | 42 |
| 1992 | 6004 | 9249 | 15253 | 61 |
| 1993 | 6872 | 4759 | 11631 | 41 |
| 1994 | 5901 | 3455 | 9356 | 37 |
| 1995 | 6076 | 5771 | 11847 | 49 |
| 1996 | 7156 | 7940 | 15096 | 53 |
| 1997 | 6285 | 5251 | 11536 | 46 |
| 1998 | 4631 | 9216 | 13847 | 67 |
| 1999 | 4613 | 3975 | 8588 | 46 |
| 2000 | 3010 | 13285 | 16295 | 82 |
| 2001 | 2438 | 4263 | 6701 | 64 |
| 2002 | 1709 | 2851 | 4560 | 63 |
| 2003 | 1331 | 1984 | 3316 | 60 |
| 2004 | 798 | 2887 | 3686 | 78 |
| 2005 | 335 | 972 | 1307 | 74 |
| 2006 | 378 | 746 | 1124 | 66 |
| 2007 | 481 | 366 | 847 | 43 |


| Year | Landings | Discards | Total | Discard rate (\%) |
| :---: | :---: | :---: | :---: | :---: |
| 2008 | 441 | 156 | 598 | 26 |
| 2009 | 480 | 826 | 1305 | 63 |
| 2010 | 345 | 1091 | 1436 | 76 |
| 2011 | 231 | 630 | 861 | 73 |
| 2012 | 300 | 742 | 1042 | 71 |
| 2013 | 215 | 1172 | 1387 | 85 |
| 2014 | 181 | 745 | 926 | 80 |
| 2015 | 221 | 1458 | 1679 | 87 |
| 2016 | 227 | 1040 | 1266 | 82 |
| 2017 | 168 | 1331 | 1498 | 89 |
| 2018 | 189 | 666 | 855 | 78 |
| 2019 | 484 | 960 | 1444 | 66 |
| 2020 | 541 | 834 | 1375 | 61 |
| 2021 | 852 | 261 | 1114 | 23 |
| Min | 168 | 156 | 598 | 15 |
| Mean | 4524 | 3650 | 8174 | 55 |
| Max | 16458 | 13285 | 23462 | 89 |

Table 35.3. Whiting in Division 6.a. Landings-at-age (thousands).

|  | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 0 | 3593 | 24395 | 11297 | 4611 | 1518 | 452 | 201 |
| 1982 | 0 | 2991 | 5783 | 29094 | 6821 | 2043 | 803 | 348 |
| 1983 | 0 | 3418 | 7094 | 8040 | 22757 | 6070 | 1439 | 540 |
| 1984 | 0 | 7209 | 12765 | 8221 | 4387 | 14825 | 1953 | 858 |
| 1985 | 0 | 4139 | 19520 | 8574 | 3351 | 1997 | 4764 | 822 |
| 1986 | 0 | 2674 | 14824 | 9770 | 2653 | 532 | 291 | 529 |
| 1987 | 0 | 6430 | 13935 | 13988 | 5442 | 837 | 330 | 259 |
| 1988 | 0 | 1842 | 20587 | 9638 | 6168 | 1949 | 290 | 207 |
| 1989 | 0 | 2529 | 5887 | 11889 | 4767 | 1266 | 468 | 71 |
| 1990 | 0 | 3203 | 8028 | 2393 | 4009 | 1326 | 204 | 37 |
| 1991 | 0 | 3294 | 8826 | 10046 | 1208 | 1391 | 286 | 51 |
| 1992 | 0 | 2695 | 9440 | 4473 | 4782 | 396 | 373 | 106 |
| 1993 | 0 | 1051 | 10179 | 6293 | 2673 | 2738 | 163 | 147 |
| 1994 | 0 | 909 | 4889 | 9158 | 3607 | 712 | 715 | 69 |
| 1995 | 0 | 215 | 4322 | 6516 | 5654 | 1397 | 376 | 282 |
| 1996 | 0 | 990 | 5410 | 7675 | 5052 | 2461 | 583 | 157 |
| 1997 | 0 | 877 | 3658 | 8514 | 4316 | 1441 | 338 | 106 |
| 1998 | 0 | 840 | 3504 | 4277 | 3698 | 1442 | 338 | 288 |
| 1999 | 0 | 1013 | 6131 | 4546 | 2040 | 1774 | 355 | 112 |
| 2000 | 0 | 484 | 2952 | 4211 | 1570 | 485 | 328 | 89 |
| 2001 | 0 | 461 | 3271 | 2630 | 1567 | 401 | 131 | 16 |
| 2002 | 0 | 62 | 1624 | 3018 | 799 | 227 | 23 | 13 |
| 2003 | 0 | 98 | 652 | 1309 | 1481 | 414 | 93 | 2 |
| 2004 | 0 | 49 | 699 | 544 | 517 | 620 | 74 | 33 |
| 2005 | 0 | 26 | 273 | 460 | 145 | 107 | 49 | 5 |
| 2006 | 0 | 83 | 135 | 386 | 276 | 67 | 86 | 25 |
| 2007 | 0 | 193 | 190 | 294 | 361 | 152 | 31 | 53 |
| 2008 | 0 | 3 | 277 | 387 | 335 | 150 | 54 | 25 |
| 2009 | 0 | 108 | 255 | 258 | 417 | 107 | 49 | 14 |
| 2010 | 0 | 50 | 81 | 150 | 148 | 141 | 43 | 52 |
| 2011 | 0 | 0 | 256 | 144 | 94 | 27 | 26 | 8 |


| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 2012 | 0 | 13 | 39 | 374 | 203 | 53 | 16 | 9 |
| 2013 | 0 | 4 | 41 | 76 | 269 | 74 | 19 | 6 |
| 2014 | 0 | 13 | 26 | 130 | 101 | 101 | 23 | 11 |
| 2015 | 0 | 7 | 74 | 56 | 157 | 71 | 73 | 30 |
| 2016 | 0 | 19 | 93 | 147 | 77 | 86 | 19 | 28 |
| 2017 | 0 | 17 | 37 | 167 | 69 | 52 | 39 | 10 |
| 2018 | 0 | 0 | 73 | 89 | 199 | 60 | 8 | 8 |
| 2019 | 0 | 23 | 54 | 427 | 255 | 258 | 48 | 5 |
| 2020 | 0 | 7 | 309 | 258 | 310 | 156 | 39 | 3 |
| 2021 | 0 | 0 | 318 | 674 | 375 | 366 | 67 | 37 |

Table 35.4. Whiting in Division 6.a. Discards-at-age (thousands). Previous discard estimates for the years 1978-2003 were replaced by those estimated by Millar and Fryer (2005).

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | NA | 1128 | 10415 | 1397 | 201 | 27 | 12 | 0 |
| 1982 | NA | 19511 | 3421 | 12683 | 1197 | 187 | 4 | 0 |
| 1983 | NA | 21690 | 6748 | 2909 | 5372 | 158 | 8 | 0 |
| 1984 | NA | 34330 | 2400 | 909 | 371 | 811 | 73 | 1 |
| 1985 | NA | 17615 | 9858 | 3273 | 672 | 205 | 363 | 40 |
| 1986 | NA | 6159 | 9823 | 1962 | 185 | 1 | 0 | 10 |
| 1987 | NA | 97611 | 17427 | 1763 | 154 | 0 | 0 | 0 |
| 1988 | NA | 28057 | 38019 | 2239 | 467 | 11 | 0 | 0 |
| 1989 | NA | 31079 | 5598 | 8570 | 223 | 13 | 5 | 0 |
| 1990 | NA | 20952 | 11176 | 71 | 23 | 3 | 0 | 0 |
| 1991 | NA | 23211 | 7540 | 7355 | 266 | 236 | 56 | 0 |
| 1992 | NA | 50665 | 16729 | 2810 | 954 | 0 | 0 | 0 |
| 1993 | NA | 14057 | 11139 | 2903 | 588 | 431 | 0 | 1 |
| 1994 | NA | 12700 | 6859 | 3872 | 1152 | 189 | 150 | 4 |
| 1995 | NA | 21974 | 21786 | 3416 | 484 | 7 | 1 | 1 |
| 1996 | NA | 33621 | 18625 | 5086 | 1535 | 13 | 1 | 20 |
| 1997 | NA | 22422 | 9632 | 3806 | 540 | 71 | 2 | 1 |
| 1998 | NA | 53742 | 16058 | 3553 | 847 | 177 | 31 | 8 |
| 1999 | NA | 7928 | 17097 | 1402 | 503 | 275 | 44 | 0 |
| 2000 | NA | 158913 | 5254 | 2238 | 154 | 16 | 41 | 0 |
| 2001 | NA | 5666 | 23084 | 715 | 172 | 0 | 0 | 0 |
| 2002 | NA | 11055 | 8531 | 2428 | 415 | 175 | 9 | 3 |
| 2003 | 5678 | 9448 | 2489 | 1775 | 375 | 25 | 7 | 1 |
| 2004 | 10577 | 14941 | 5095 | 1011 | 660 | 125 | 4 | 2 |
| 2005 | 7873 | 3246 | 2298 | 769 | 60 | 22 | 8 | 4 |
| 2006 | 5866 | 4691 | 528 | 637 | 169 | 29 | 6 | 2 |
| 2007 | 1259 | 1016 | 966 | 283 | 88 | 38 | 3 | 0 |
| 2008 | 840 | 630 | 144 | 114 | 31 | 37 | 4 | 0 |
| 2009 | 9685 | 6880 | 114 | 66 | 44 | 15 | 4 | 0 |
| 2010 | 5903 | 17678 | 1581 | 264 | 37 | 54 | 6 | 16 |


| Year | Age |  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2011 | 13306 | 2047 | 998 | 122 | 7 | 2 | 0 | $\mathbf{7}+$ |
| 2012 | 1434 | 7810 | 429 | 547 | 94 | 19 | 1 | 0 |
| 2013 | 3188 | 16415 | 1578 | 172 | 255 | 8 | 2 | 0 |
| 2014 | 6261 | 9831 | 51 | 55 | 27 | 30 | 8 | 2 |
| 2015 | 17740 | 7930 | 909 | 287 | 112 | 18 | 17 | 0 |
| 2016 | 3745 | 5506 | 1910 | 268 | 16 | 12 | 4 | 2 |
| 2017 | 8518 | 7563 | 788 | 889 | 65 | 160 | 2 | 0 |
| 2018 | 1777 | 2371 | 962 | 469 | 276 | 21 | 5 | 0 |
| 2019 | 2188 | 10379 | 526 | 413 | 232 | 34 | 0 | 0 |
| 2021 | 2579 | 156 | 467 | 80 | 59 | 44 | 89 | 0 |

Table 35.5. Whiting in Division 6.a. Total catch-at-age (thousands).

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | NA | 4721 | 34810 | 12694 | 4812 | 1545 | 464 | 201 |
| 1982 | NA | 22502 | 9204 | 41777 | 8018 | 2230 | 807 | 348 |
| 1983 | NA | 25108 | 13842 | 10949 | 28129 | 6228 | 1447 | 540 |
| 1984 | NA | 41539 | 15165 | 9130 | 4758 | 15636 | 2026 | 859 |
| 1985 | NA | 21754 | 29378 | 11847 | 4023 | 2202 | 5127 | 862 |
| 1986 | NA | 8833 | 24647 | 11732 | 2838 | 533 | 291 | 539 |
| 1987 | NA | 104041 | 31362 | 15751 | 5596 | 837 | 330 | 259 |
| 1988 | NA | 29899 | 58606 | 11877 | 6635 | 1960 | 290 | 207 |
| 1989 | NA | 33608 | 11485 | 20459 | 4990 | 1279 | 473 | 71 |
| 1990 | NA | 24155 | 19204 | 2464 | 4032 | 1329 | 204 | 37 |
| 1991 | NA | 26505 | 16366 | 17401 | 1474 | 1627 | 342 | 51 |
| 1992 | NA | 53360 | 26169 | 7283 | 5736 | 396 | 373 | 106 |
| 1993 | NA | 15108 | 21318 | 9196 | 3261 | 3169 | 163 | 148 |
| 1994 | NA | 13609 | 11748 | 13030 | 4759 | 901 | 865 | 73 |
| 1995 | NA | 22189 | 26108 | 9932 | 6138 | 1404 | 377 | 283 |
| 1996 | NA | 34611 | 24035 | 12761 | 6587 | 2474 | 584 | 177 |
| 1997 | NA | 23299 | 13290 | 12320 | 4856 | 1512 | 340 | 107 |
| 1998 | NA | 54582 | 19562 | 7830 | 4545 | 1619 | 369 | 296 |
| 1999 | NA | 8941 | 23228 | 5948 | 2543 | 2049 | 399 | 112 |
| 2000 | NA | 159397 | 8206 | 6449 | 1724 | 501 | 369 | 89 |
| 2001 | NA | 6127 | 26355 | 3345 | 1739 | 401 | 131 | 16 |
| 2002 | NA | 11117 | 10155 | 5446 | 1214 | 402 | 32 | 16 |
| 2003 | 5678 | 9546 | 3141 | 3083 | 1856 | 439 | 100 | 3 |
| 2004 | 10577 | 14990 | 5794 | 1556 | 1176 | 745 | 78 | 35 |
| 2005 | 7873 | 3272 | 2571 | 1229 | 205 | 129 | 57 | 10 |
| 2006 | 5866 | 4773 | 663 | 1023 | 445 | 96 | 93 | 27 |
| 2007 | 1259 | 1209 | 1156 | 578 | 449 | 190 | 33 | 53 |
| 2008 | 840 | 632 | 421 | 500 | 366 | 187 | 58 | 25 |
| 2009 | 9685 | 6988 | 370 | 324 | 462 | 123 | 53 | 14 |
| 2010 | 5903 | 17729 | 1662 | 414 | 185 | 196 | 49 | 68 |
| 2011 | 13306 | 2048 | 1254 | 267 | 101 | 29 | 26 | 8 |


|  | Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2012 | 1434 | 7823 | 469 | 920 | 298 | 72 | 17 | 9 |
| 2013 | 3188 | 16419 | 1619 | 247 | 523 | 82 | 21 | 7 |
| 2014 | 6261 | 9844 | 77 | 185 | 127 | 130 | 31 | 14 |
| 2015 | 17740 | 7937 | 983 | 343 | 269 | 90 | 90 | 30 |
| 2016 | 3745 | 5525 | 2003 | 415 | 92 | 98 | 23 | 30 |
| 2017 | 8518 | 7580 | 825 | 1056 | 134 | 212 | 41 | 10 |
| 2018 | 1777 | 2371 | 1035 | 557 | 475 | 81 | 13 | 8 |
| 2019 | 2188 | 10402 | 580 | 840 | 486 | 293 | 48 | 5 |
| 2020 | NA | 23488 | 1116 | 317 | 339 | 166 | 42 | 3 |
| 2021 | NA | NA | 785 | 754 | 419 | 456 | 67 | 37 |

Table 35.6. Whiting in Division 6.a. Mean weight-at-age (kg) in landings.

| Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0* | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 0.035 | 0.192 | 0.228 | 0.289 | 0.382 | 0.409 | 0.409 | 0.547 |
| 1982 | 0.035 | 0.184 | 0.22 | 0.276 | 0.352 | 0.505 | 0.513 | 0.526 |
| 1983 | 0.035 | 0.216 | 0.249 | 0.28 | 0.34 | 0.409 | 0.494 | 0.51 |
| 1984 | 0.035 | 0.216 | 0.259 | 0.313 | 0.371 | 0.412 | 0.458 | 0.458 |
| 1985 | 0.035 | 0.185 | 0.238 | 0.306 | 0.402 | 0.43 | 0.461 | 0.538 |
| 1986 | 0.035 | 0.174 | 0.236 | 0.294 | 0.365 | 0.468 | 0.482 | 0.499 |
| 1987 | 0.035 | 0.188 | 0.237 | 0.304 | 0.373 | 0.511 | 0.52 | 0.576 |
| 1988 | 0.035 | 0.176 | 0.215 | 0.301 | 0.4 | 0.483 | 0.567 | 0.6 |
| 1989 | 0.035 | 0.171 | 0.22 | 0.279 | 0.348 | 0.459 | 0.425 | 0.555 |
| 1990 | 0.035 | 0.225 | 0.251 | 0.324 | 0.359 | 0.417 | 0.582 | 0.543 |
| 1991 | 0.035 | 0.199 | 0.22 | 0.291 | 0.354 | 0.391 | 0.442 | 0.761 |
| 1992 | 0.035 | 0.193 | 0.23 | 0.288 | 0.349 | 0.388 | 0.397 | 0.51 |
| 1993 | 0.035 | 0.186 | 0.242 | 0.314 | 0.361 | 0.412 | 0.452 | 0.474 |
| 1994 | 0.035 | 0.161 | 0.217 | 0.29 | 0.371 | 0.451 | 0.482 | 0.483 |
| 1995 | 0.035 | 0.19 | 0.225 | 0.296 | 0.381 | 0.469 | 0.473 | 0.528 |
| 1996 | 0.035 | 0.195 | 0.245 | 0.288 | 0.365 | 0.483 | 0.526 | 0.569 |
| 1997 | 0.035 | 0.198 | 0.245 | 0.297 | 0.384 | 0.522 | 0.629 | 0.661 |
| 1998 | 0.035 | 0.215 | 0.236 | 0.301 | 0.364 | 0.438 | 0.5 | 0.646 |
| 1999 | 0.035 | 0.181 | 0.225 | 0.28 | 0.365 | 0.44 | 0.524 | 0.594 |
| 2000 | 0.035 | 0.205 | 0.241 | 0.298 | 0.336 | 0.419 | 0.488 | 0.617 |
| 2001 | 0.035 | 0.173 | 0.234 | 0.303 | 0.37 | 0.395 | 0.376 | 0.595 |
| 2002 | 0.035 | 0.213 | 0.257 | 0.304 | 0.363 | 0.464 | 0.650 | 0.707 |
| 2003 | 0.035 | 0.236 | 0.272 | 0.301 | 0.373 | 0.349 | 0.409 | 0.659 |
| 2004 | 0.035 | 0.189 | 0.257 | 0.296 | 0.342 | 0.376 | 0.378 | 0.305 |
| 2005 | 0.035 | 0.215 | 0.253 | 0.297 | 0.366 | 0.426 | 0.455 | 0.383 |
| 2006 | 0.035 | 0.221 | 0.290 | 0.321 | 0.395 | 0.452 | 0.496 | 0.574 |
| 2007 | 0.035 | 0.215 | 0.289 | 0.356 | 0.416 | 0.497 | 0.598 | 0.667 |
| 2008 | 0.035 | 0.285 | 0.245 | 0.319 | 0.379 | 0.516 | 0.534 | 0.652 |
| 2009 | 0.035 | 0.288 | 0.317 | 0.406 | 0.446 | 0.439 | 0.444 | 0.603 |
| 2010 | 0.035 | 0.286 | 0.353 | 0.436 | 0.540 | 0.647 | 0.654 | 0.575 |
| 2011 | 0.035 | 0.201 | 0.356 | 0.396 | 0.502 | 0.571 | 0.578 | 0.370 |


| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0* | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 2012 | 0.035 | 0.320 | 0.300 | 0.374 | 0.504 | 0.594 | 0.665 | 0.482 |
| 2013 | 0.035 | 0.225 | 0.325 | 0.355 | 0.441 | 0.546 | 0.597 | 0.770 |
| 2014 | 0.035 | 0.248 | 0.295 | 0.375 | 0.457 | 0.528 | 0.641 | 0.678 |
| 2015 | 0.035 | 0.261 | 0.347 | 0.447 | 0.468 | 0.508 | 0.596 | 0.600 |
| 2016 | 0.035 | 0.137 | 0.325 | 0.483 | 0.509 | 0.606 | 0.676 | 0.664 |
| 2017 | 0.035 | 0.340 | 0.352 | 0.413 | 0.546 | 0.497 | 0.510 | 0.684 |
| 2018 | 0.035 | 0.173 | 0.407 | 0.396 | 0.435 | 0.520 | 0.472 | 0.564 |
| 2019 | 0.035 | 0.244 | 0.288 | 0.415 | 0.506 | 0.529 | 0.698 | 0.879 |
| 2020 | 0.035 | 0.235 | 0.406 | 0.482 | 0.551 | 0.597 | 0.657 | 1.058 |
| 2021 | 0.035 | 0.217** | 0.301 | 0.418 | 0.585 | 0.577 | 0.620 | 0.658 |

* For age 0, mean weights-are assumed to be the average for discards in 2003-2019.
** For age 1 in 2021, mean weight-is assumed to be the average for age 1 in 2018-2020.

Table 35.7. Whiting in Division 6.a. Mean weight-at-age (kg) in discards.

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 0.035* | 0.108 | 0.16 | 0.195 | 0.298 | 0.286 | 0.295 | NA |
| 1982 | 0.035* | 0.096 | 0.18 | 0.209 | 0.243 | 0.283 | 0.44 | NA |
| 1983 | 0.035* | 0.141 | 0.186 | 0.228 | 0.237 | 0.267 | 0.267 | NA |
| 1984 | 0.035* | 0.087 | 0.199 | 0.246 | 0.26 | 0.259 | 0.303 | 0.227 |
| 1985 | 0.035* | 0.102 | 0.191 | 0.237 | 0.286 | 0.326 | 0.312 | 0.316 |
| 1986 | 0.035* | 0.092 | 0.17 | 0.196 | 0.245 | 0.258 | 0.33 | 0.263 |
| 1987 | 0.035* | 0.085 | 0.182 | 0.233 | 0.249 | 0.225 | NA | NA |
| 1988 | 0.035* | 0.076 | 0.143 | 0.203 | 0.227 | 0.262 | NA | NA |
| 1989 | 0.035* | 0.099 | 0.177 | 0.205 | 0.209 | 0.294 | 0.305 | NA |
| 1990 | 0.035* | 0.124 | 0.171 | 0.214 | 0.219 | 0.237 | 0.264 | NA |
| 1991 | 0.035* | 0.085 | 0.169 | 0.205 | 0.223 | 0.226 | 0.281 | NA |
| 1992 | 0.035* | 0.109 | 0.173 | 0.219 | 0.227 | NA | NA | NA |
| 1993 | 0.035* | 0.118 | 0.197 | 0.225 | 0.242 | 0.256 | NA | 0.436 |
| 1994 | 0.035* | 0.087 | 0.157 | 0.22 | 0.283 | 0.297 | 0.253 | 0.299 |
| 1995 | 0.035* | 0.075 | 0.154 | 0.189 | 0.246 | 0.278 | 0.597 | 0.493 |
| 1996 | 0.035* | 0.095 | 0.18 | 0.203 | 0.229 | 0.302 | 0.421 | 0.26 |
| 1997 | 0.035* | 0.112 | 0.182 | 0.221 | 0.235 | 0.243 | 0.422 | 0.819 |
| 1998 | 0.035* | 0.098 | 0.179 | 0.225 | 0.254 | 0.282 | 0.264 | 0.245 |
| 1999 | 0.035* | 0.077 | 0.168 | 0.217 | 0.205 | 0.266 | 0.268 | NA |
| 2000 | 0.035* | 0.075 | 0.164 | 0.203 | 0.233 | 0.282 | 0.250 | NA |
| 2001 | 0.035* | 0.094 | 0.154 | 0.196 | 0.203 | 0.381 | 0.000 | NA |
| 2002 | 0.035* | 0.073 | 0.162 | 0.212 | 0.245 | 0.240 | 0.295 | 0.276 |
| 2003 | 0.051 | 0.091 | 0.161 | 0.193 | 0.243 | 0.209 | 0.291 | 0.278 |
| 2004 | 0.020 | 0.091 | 0.178 | 0.223 | 0.233 | 0.302 | 0.343 | 0.282 |
| 2005 | 0.028 | 0.074 | 0.145 | 0.207 | 0.188 | 0.302 | 0.289 | 0.368 |
| 2006 | 0.037 | 0.047 | 0.195 | 0.233 | 0.285 | 0.311 | 0.494 | 0.361 |
| 2007 | 0.042 | 0.064 | 0.157 | 0.232 | 0.223 | 0.231 | 0.787 | 0.266 |
| 2008 | 0.019 | 0.076 | 0.211 | 0.305 | 0.350 | 0.423 | 0.233 | 0.289 |
| 2009 | 0.043 | 0.051 | 0.283 | 0.227 | 0.262 | 0.250 | 0.248 | NA |
| 2010 | 0.018 | 0.040 | 0.119 | 0.239 | 0.360 | 0.360 | 0.382 | 0.224 |
| 2011 | 0.029 | 0.034 | 0.136 | 0.307 | 0.256 | 0.228 | NA | NA |


| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 2012 | 0.042 | 0.057 | 0.152 | 0.292 | 0.362 | 0.356 | 0.386 | NA |
| 2013 | 0.027 | 0.041 | 0.209 | 0.229 | 0.358 | 0.385 | 0.299 | 0.371 |
| 2014 | 0.040 | 0.045 | 0.182 | 0.289 | 0.362 | 0.427 | 0.422 | 0.757 |
| 2015 | 0.035 | 0.072 | 0.171 | 0.212 | 0.336 | 0.316 | 0.427 | NA |
| 2016 | 0.050 | 0.068 | 0.206 | 0.276 | 0.292 | 0.304 | 0.261 | 0.367 |
| 2017 | 0.033 | 0.066 | 0.197 | 0.351 | 0.409 | 0.331 | 0.881 | NA |
| 2018 | 0.054 | 0.067 | 0.184 | 0.250 | 0.307 | 0.414 | 1.107 | NA |
| 2019 | 0.029 | 0.055 | 0.199 | 0.267 | 0.278 | 0.436 | 0.489 | NA |
| 2020 | 0.035* | 0.028 | 0.163 | 0.254 | 0.313 | 0.286 | 0.255 | NA |
| 2021 | 0.029 | 0.143 | 0.236 | 0.279 | 0.270 | 0.236 | NA | NA |

* For age 0 in 1981-2002 and 2020, mean weights-are assumed to be the average for 2003-2019.

Table 35.8. Whiting in Division 6.a. Mean weight-at-age (kg) in total catch.

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 0.035* | 0.172 | 0.208 | 0.279 | 0.378 | 0.407 | 0.406 | 0.547 |
| 1982 | 0.035* | 0.108 | 0.205 | 0.256 | 0.336 | 0.486 | 0.513 | 0.526 |
| 1983 | 0.035* | 0.151 | 0.218 | 0.266 | 0.320 | 0.405 | 0.493 | 0.510 |
| 1984 | 0.035* | 0.109 | 0.250 | 0.306 | 0.362 | 0.404 | 0.452 | 0.458 |
| 1985 | 0.035* | 0.118 | 0.222 | 0.287 | 0.383 | 0.420 | 0.450 | 0.528 |
| 1986 | 0.035* | 0.117 | 0.210 | 0.278 | 0.357 | 0.468 | 0.482 | 0.495 |
| 1987 | 0.035* | 0.091 | 0.206 | 0.296 | 0.370 | 0.511 | 0.520 | 0.576 |
| 1988 | 0.035* | 0.082 | 0.168 | 0.283 | 0.388 | 0.482 | 0.567 | 0.600 |
| 1989 | 0.035* | 0.104 | 0.199 | 0.248 | 0.342 | 0.457 | 0.424 | 0.555 |
| 1990 | 0.035* | 0.137 | 0.204 | 0.321 | 0.358 | 0.417 | 0.582 | 0.543 |
| 1991 | 0.035* | 0.099 | 0.197 | 0.255 | 0.330 | 0.367 | 0.416 | 0.761 |
| 1992 | 0.035* | 0.113 | 0.194 | 0.261 | 0.329 | 0.388 | 0.397 | 0.510 |
| 1993 | 0.035* | 0.123 | 0.218 | 0.286 | 0.340 | 0.391 | 0.452 | 0.474 |
| 1994 | 0.035* | 0.092 | 0.182 | 0.269 | 0.350 | 0.419 | 0.442 | 0.473 |
| 1995 | 0.035* | 0.076 | 0.166 | 0.259 | 0.370 | 0.468 | 0.473 | 0.528 |
| 1996 | 0.035* | 0.098 | 0.195 | 0.254 | 0.333 | 0.482 | 0.526 | 0.534 |
| 1997 | 0.035* | 0.115 | 0.199 | 0.274 | 0.367 | 0.509 | 0.628 | 0.662 |
| 1998 | 0.035* | 0.100 | 0.189 | 0.267 | 0.344 | 0.421 | 0.480 | 0.635 |
| 1999 | 0.035* | 0.089 | 0.183 | 0.265 | 0.333 | 0.417 | 0.496 | 0.594 |
| 2000 | 0.035* | 0.075 | 0.192 | 0.265 | 0.327 | 0.415 | 0.462 | 0.617 |
| 2001 | 0.035* | 0.100 | 0.164 | 0.280 | 0.353 | 0.395 | 0.376 | 0.595 |
| 2002 | 0.035* | 0.074 | 0.177 | 0.263 | 0.323 | 0.366 | 0.550 | 0.626 |
| 2003 | 0.051 | 0.092 | 0.184 | 0.239 | 0.347 | 0.341 | 0.401 | 0.516 |
| 2004 | 0.020 | 0.091 | 0.188 | 0.249 | 0.281 | 0.364 | 0.377 | 0.304 |
| 2005 | 0.028 | 0.075 | 0.156 | 0.241 | 0.313 | 0.405 | 0.432 | 0.376 |
| 2006 | 0.037 | 0.050 | 0.214 | 0.266 | 0.353 | 0.410 | 0.495 | 0.557 |
| 2007 | 0.042 | 0.088 | 0.179 | 0.295 | 0.378 | 0.444 | 0.613 | 0.666 |
| 2008 | 0.019 | 0.077 | 0.233 | 0.316 | 0.376 | 0.498 | 0.514 | 0.648 |
| 2009 | 0.043 | 0.054 | 0.307 | 0.369 | 0.429 | 0.415 | 0.430 | 0.603 |
| 2010 | 0.018 | 0.040 | 0.130 | 0.311 | 0.504 | 0.567 | 0.622 | 0.492 |
| 2011 | 0.029 | 0.034 | 0.181 | 0.355 | 0.485 | 0.546 | 0.578 | 0.370 |


| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 2012 | 0.042 | 0.057 | 0.164 | 0.325 | 0.459 | 0.531 | 0.643 | 0.482 |
| 2013 | 0.027 | 0.041 | 0.212 | 0.268 | 0.401 | 0.530 | 0.571 | 0.679 |
| 2014 | 0.040 | 0.045 | 0.220 | 0.349 | 0.437 | 0.505 | 0.581 | 0.694 |
| 2015 | 0.035 | 0.072 | 0.185 | 0.250 | 0.413 | 0.469 | 0.565 | 0.600 |
| 2016 | 0.050 | 0.068 | 0.211 | 0.349 | 0.472 | 0.568 | 0.601 | 0.649 |
| 2017 | 0.033 | 0.066 | 0.204 | 0.361 | 0.480 | 0.372 | 0.524 | 0.684 |
| 2018 | 0.054 | 0.067 | 0.199 | 0.273 | 0.361 | 0.492 | 0.731 | 0.564 |
| 2019 | 0.029 | 0.055 | 0.207 | 0.342 | 0.397 | 0.518 | 0.697 | 0.879 |
| 2020 | 0.035* | 0.028 | 0.230 | 0.439 | 0.531 | 0.579 | 0.625 | 1.058 |
| 2021 | 0.029 | 0.143 | 0.262 | 0.403 | 0.552 | 0.510 | 0.620 | 0.658 |

* For age 0 in 1981-2002 and 2020, mean weights-are assumed to be the average for 2003-2019.

Table 35.9. Whiting in Division 6.a. Mean weight-at-age (kg) in stock. These are smoothed estimates for use in SSB calculation.

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 0.037 | 0.048 | 0.153 | 0.273 | 0.373 | 0.48 | 0.552 | 0.648 |
| 1982 | 0.037 | 0.048 | 0.153 | 0.273 | 0.373 | 0.48 | 0.552 | 0.648 |
| 1983 | 0.037 | 0.048 | 0.153 | 0.273 | 0.373 | 0.48 | 0.552 | 0.648 |
| 1984 | 0.037 | 0.048 | 0.153 | 0.273 | 0.373 | 0.48 | 0.552 | 0.648 |
| 1985 | 0.037 | 0.048 | 0.153 | 0.273 | 0.373 | 0.48 | 0.552 | 0.648 |
| 1986 | 0.037 | 0.048 | 0.153 | 0.27 | 0.369 | 0.477 | 0.55 | 0.642 |
| 1987 | 0.037 | 0.048 | 0.153 | 0.267 | 0.365 | 0.472 | 0.547 | 0.637 |
| 1988 | 0.037 | 0.048 | 0.152 | 0.264 | 0.361 | 0.464 | 0.545 | 0.631 |
| 1989 | 0.036 | 0.048 | 0.152 | 0.261 | 0.358 | 0.454 | 0.543 | 0.624 |
| 1990 | 0.036 | 0.048 | 0.151 | 0.259 | 0.356 | 0.442 | 0.54 | 0.617 |
| 1991 | 0.036 | 0.049 | 0.15 | 0.256 | 0.354 | 0.433 | 0.538 | 0.61 |
| 1992 | 0.036 | 0.049 | 0.148 | 0.254 | 0.353 | 0.428 | 0.536 | 0.603 |
| 1993 | 0.036 | 0.049 | 0.146 | 0.252 | 0.352 | 0.431 | 0.533 | 0.596 |
| 1994 | 0.035 | 0.049 | 0.145 | 0.25 | 0.35 | 0.439 | 0.531 | 0.589 |
| 1995 | 0.035 | 0.049 | 0.143 | 0.249 | 0.349 | 0.449 | 0.53 | 0.582 |
| 1996 | 0.035 | 0.049 | 0.141 | 0.247 | 0.347 | 0.457 | 0.528 | 0.576 |
| 1997 | 0.035 | 0.05 | 0.14 | 0.246 | 0.344 | 0.459 | 0.527 | 0.57 |
| 1998 | 0.035 | 0.05 | 0.138 | 0.245 | 0.341 | 0.453 | 0.527 | 0.565 |
| 1999 | 0.034 | 0.05 | 0.138 | 0.245 | 0.337 | 0.44 | 0.526 | 0.56 |
| 2000 | 0.034 | 0.05 | 0.137 | 0.245 | 0.334 | 0.424 | 0.526 | 0.556 |
| 2001 | 0.034 | 0.05 | 0.137 | 0.246 | 0.331 | 0.409 | 0.527 | 0.553 |
| 2002 | 0.034 | 0.05 | 0.137 | 0.248 | 0.33 | 0.398 | 0.528 | 0.551 |
| 2003 | 0.034 | 0.05 | 0.138 | 0.25 | 0.332 | 0.395 | 0.531 | 0.551 |
| 2004 | 0.034 | 0.051 | 0.139 | 0.254 | 0.337 | 0.401 | 0.534 | 0.552 |
| 2005 | 0.033 | 0.051 | 0.141 | 0.26 | 0.346 | 0.415 | 0.538 | 0.555 |
| 2006 | 0.033 | 0.051 | 0.143 | 0.266 | 0.359 | 0.435 | 0.543 | 0.56 |
| 2007 | 0.033 | 0.051 | 0.146 | 0.273 | 0.374 | 0.458 | 0.549 | 0.566 |
| 2008 | 0.033 | 0.051 | 0.148 | 0.281 | 0.392 | 0.483 | 0.556 | 0.574 |
| 2009 | 0.033 | 0.051 | 0.151 | 0.289 | 0.409 | 0.507 | 0.564 | 0.584 |
| 2010 | 0.033 | 0.051 | 0.154 | 0.296 | 0.425 | 0.529 | 0.571 | 0.594 |


|  | Age | Pear | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 0.034 | 0.052 | 0.156 | 0.302 | 0.438 | 0.547 | 0.579 | 0.606 |
| 2012 | 0.034 | 0.052 | 0.158 | 0.307 | 0.448 | 0.558 | 0.587 | 0.619 |
| 2013 | 0.034 | 0.052 | 0.16 | 0.31 | 0.454 | 0.563 | 0.595 | 0.633 |
| 2014 | 0.034 | 0.052 | 0.162 | 0.314 | 0.456 | 0.56 | 0.602 | 0.648 |
| 2015 | 0.034 | 0.052 | 0.163 | 0.316 | 0.456 | 0.553 | 0.609 | 0.663 |
| 2016 | 0.035 | 0.052 | 0.164 | 0.319 | 0.455 | 0.543 | 0.616 | 0.678 |
| 2018 | 0.035 | 0.052 | 0.164 | 0.322 | 0.454 | 0.535 | 0.622 | 0.694 |
| 2019 | 0.036 | 0.053 | 0.163 | 0.327 | 0.454 | 0.529 | 0.635 | 0.725 |
| 2020 | 0.036 | 0.053 | 0.162 | 0.33 | 0.455 | 0.532 | 0.641 | 0.74 |
| 2021 | 0.036 | 0.053 | 0.161 | 0.333 | 0.457 | 0.538 | 0.647 | 0.755 |

Table 35.10. Whiting in Division 6.a. Natural mortality.

| Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 1.052 | 1.036 | 0.713 | 0.593 | 0.539 | 0.501 | 0.478 | 0.459 |
| 1982 | 1.052 | 1.036 | 0.713 | 0.593 | 0.539 | 0.501 | 0.478 | 0.459 |
| 1983 | 1.052 | 1.036 | 0.713 | 0.593 | 0.539 | 0.501 | 0.478 | 0.459 |
| 1984 | 1.052 | 1.036 | 0.713 | 0.593 | 0.539 | 0.501 | 0.478 | 0.459 |
| 1985 | 1.052 | 1.036 | 0.713 | 0.593 | 0.539 | 0.501 | 0.478 | 0.459 |
| 1986 | 1.054 | 1.015 | 0.710 | 0.594 | 0.541 | 0.501 | 0.479 | 0.460 |
| 1987 | 1.055 | 0.996 | 0.708 | 0.595 | 0.543 | 0.503 | 0.480 | 0.461 |
| 1988 | 1.057 | 0.979 | 0.705 | 0.597 | 0.544 | 0.505 | 0.482 | 0.463 |
| 1989 | 1.058 | 0.963 | 0.702 | 0.598 | 0.545 | 0.509 | 0.483 | 0.464 |
| 1990 | 1.060 | 0.948 | 0.699 | 0.599 | 0.546 | 0.513 | 0.484 | 0.465 |
| 1991 | 1.062 | 0.935 | 0.696 | 0.600 | 0.547 | 0.517 | 0.485 | 0.467 |
| 1992 | 1.063 | 0.921 | 0.694 | 0.601 | 0.548 | 0.519 | 0.487 | 0.469 |
| 1993 | 1.065 | 0.909 | 0.691 | 0.601 | 0.548 | 0.518 | 0.488 | 0.470 |
| 1994 | 1.066 | 0.897 | 0.689 | 0.601 | 0.548 | 0.515 | 0.489 | 0.472 |
| 1995 | 1.068 | 0.885 | 0.686 | 0.601 | 0.548 | 0.511 | 0.490 | 0.473 |
| 1996 | 1.070 | 0.874 | 0.683 | 0.600 | 0.547 | 0.507 | 0.491 | 0.475 |
| 1997 | 1.071 | 0.865 | 0.681 | 0.599 | 0.547 | 0.506 | 0.492 | 0.476 |
| 1998 | 1.073 | 0.857 | 0.678 | 0.598 | 0.548 | 0.507 | 0.493 | 0.478 |
| 1999 | 1.075 | 0.850 | 0.675 | 0.597 | 0.548 | 0.511 | 0.494 | 0.479 |
| 2000 | 1.077 | 0.844 | 0.673 | 0.595 | 0.549 | 0.517 | 0.495 | 0.480 |
| 2001 | 1.079 | 0.840 | 0.670 | 0.593 | 0.549 | 0.523 | 0.495 | 0.481 |
| 2002 | 1.080 | 0.837 | 0.668 | 0.591 | 0.548 | 0.529 | 0.495 | 0.482 |
| 2003 | 1.082 | 0.834 | 0.666 | 0.589 | 0.547 | 0.532 | 0.494 | 0.482 |
| 2004 | 1.083 | 0.833 | 0.664 | 0.586 | 0.545 | 0.531 | 0.493 | 0.482 |
| 2005 | 1.085 | 0.832 | 0.661 | 0.583 | 0.542 | 0.528 | 0.491 | 0.482 |
| 2006 | 1.085 | 0.832 | 0.659 | 0.580 | 0.537 | 0.522 | 0.489 | 0.481 |
| 2007 | 1.086 | 0.832 | 0.657 | 0.576 | 0.532 | 0.514 | 0.486 | 0.480 |
| 2008 | 1.086 | 0.833 | 0.655 | 0.573 | 0.527 | 0.506 | 0.484 | 0.478 |
| 2009 | 1.085 | 0.833 | 0.653 | 0.570 | 0.522 | 0.497 | 0.481 | 0.476 |
| 2010 | 1.084 | 0.833 | 0.652 | 0.567 | 0.517 | 0.490 | 0.479 | 0.474 |
| 2011 | 1.083 | 0.832 | 0.650 | 0.564 | 0.513 | 0.484 | 0.476 | 0.471 |


| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 2012 | 1.081 | 0.830 | 0.648 | 0.562 | 0.510 | 0.480 | 0.474 | 0.468 |
| 2013 | 1.080 | 0.828 | 0.647 | 0.560 | 0.509 | 0.479 | 0.471 | 0.466 |
| 2014 | 1.077 | 0.825 | 0.646 | 0.558 | 0.508 | 0.479 | 0.469 | 0.463 |
| 2015 | 1.075 | 0.823 | 0.644 | 0.556 | 0.507 | 0.481 | 0.468 | 0.460 |
| 2016 | 1.073 | 0.821 | 0.643 | 0.555 | 0.507 | 0.484 | 0.466 | 0.457 |
| 2017 | 1.070 | 0.820 | 0.642 | 0.553 | 0.507 | 0.486 | 0.465 | 0.454 |
| 2018 | 1.068 | 0.818 | 0.641 | 0.552 | 0.507 | 0.488 | 0.463 | 0.451 |
| 2019 | 1.065 | 0.817 | 0.640 | 0.550 | 0.507 | 0.487 | 0.462 | 0.448 |
| 2020 | 1.062 | 0.817 | 0.639 | 0.549 | 0.506 | 0.486 | 0.461 | 0.445 |
| 2021 | 1.060 | 0.816 | 0.638 | 0.547 | 0.506 | 0.484 | 0.460 | 0.443 |

Table 35.11. Whiting in Division 6.a. Survey data made available to the WG. For the Scottish and Irish surveys, numbers are standardised to catch-rate per ten hours. The Scottish surveys from 2011 have been conducted according to the new design and ground gear.

| ScoGFS-WIBTS-Q1 - Scottish Groundfish Survey - numbers-at-age/10 h |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Effort | Age |  |  |  |  |  |  |
|  | (hours) | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1985 | 10 | 3140 | 1792 | 380 | 85 | 23 | 156 | 18 |
| 1986 | 10 | 1456 | 1525 | 403 | 68 | 10 | 9 | 10 |
| 1987 | 10 | 6938 | 1054 | 584 | 142 | 36 | 2 | 1 |
| 1988 | 10 | 567 | 3469 | 654 | 189 | 42 | 5 | 1 |
| 1989 | 10 | 910 | 505 | 586 | 237 | 48 | 3 | 0 |
| 1990 | 10 | 1818 | 571 | 122 | 216 | 61 | 4 | 1 |
| 1991 | 10 | 3203 | 276 | 299 | 22 | 39 | 9 | 1 |
| 1992 | 10 | 4777 | 1597 | 410 | 517 | 56 | 18 | 0 |
| 1993 | 10 | 5532 | 6829 | 644 | 91 | 30 | 11 | 2 |
| 1994 | 10 | 6614 | 2443 | 1487 | 174 | 56 | 15 | 6 |
| 1995 | 10 | 5598 | 2831 | 1160 | 370 | 70 | 17 | 32 |
| 1996 | 10 | 9385 | 2237 | 635 | 341 | 135 | 30 | 4 |
| 1997 | 10 | 5663 | 2444 | 1531 | 355 | 102 | 17 | 4 |
| 1998 | 10 | 9851 | 1352 | 294 | 195 | 50 | 14 | 1 |
| 1999 | 10 | 6125 | 4952 | 489 | 103 | 16 | 1 | 0 |
| 2000 | 10 | 12862 | 471 | 152 | 34 | 10 | 11 | 0 |
| 2001 | 10 | 4653 | 1955 | 242 | 41 | 8 | 1 | 1 |
| 2002 | 10 | 5542 | 1028 | 964 | 89 | 15 | 1 | 1 |
| 2003 | 10 | 6934 | 746 | 436 | 300 | 32 | 2 | 4 |
| 2004 | 10 | 5887 | 1566 | 189 | 131 | 44 | 9 | 1 |
| 2005 | 10 | 1308 | 723 | 183 | 35 | 8 | 11 | 2 |
| 2006 | 10 | 1441 | 466 | 282 | 77 | 0 | 3 | 1 |
| 2007 | 10 | 614 | 522 | 127 | 75 | 16 | 3 | 2 |
| 2008 | 10 | 593 | 127 | 77 | 26 | 8 | 3 | 0 |
| 2009 | 10 | 906 | 387 | 103 | 105 | 20 | 9 | 7 |
| 2010 | 10 | 3523 | 340 | 108 | 52 | 40 | 4 | 3 |

Table 35.11. Continued.

| UK-SCOWCGFS-Q1 - Scottish Groundfish Survey - numbers-at-age/10 h |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Index |  |  | Variance |  |  |  |  |  |  |  |  |  |  |  |
|  | Effort | Age |  |  |  |  |  |  | Age |  |  |  |  |  |  |
|  | (hours) | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 2011 | 10 | 222 | 1884 | 397 | 64 | 37 | 45 | 12 | 6431 | 150861 | 5654 | 209 | 80 | 133 | 11 |
| 2012 | 10 | 3441 | 293 | 738 | 72 | 14 | 5 | 7 | 600264 | 8104 | 18380 | 184 | 9 | 2 | 3 |
| 2013 | 10 | 552 | 1031 | 302 | 463 | 61 | 7 | 3 | 62915 | 46672 | 5056 | 15023 | 443 | 7 | 1 |
| 2014 | 10 | 5805 | 125 | 246 | 110 | 74 | 7 | 1 | 2230995 | 556 | 2133 | 657 | 333 | 2 | 0 |
| 2015 | 10 | 2545 | 760 | 285 | 259 | 65 | 58 | 8 | 144266 | 46202 | 8599 | 4562 | 305 | 352 | 10 |
| 2016 | 10 | 3226 | 3485 | 576 | 148 | 84 | 42 | 25 | 397138 | 1880448 | 28776 | 691 | 260 | 95 | 48 |
| 2017 | 10 | 4970 | 1981 | 1707 | 203 | 49 | 32 | 5 | 2335667 | 309373 | 227966 | 2958 | 172 | 99 | 3 |
| 2018 | 10 | 1960 | 1827 | 1069 | 1142 | 132 | 14 | 2 | 763992 | 330295 | 91346 | 108990 | 2138 | 70 | 0 |
| 2019 | 10 | 3231 | 666 | 577 | 191 | 99 | 25 | 0 | 345197 | 29689 | 21447 | 1786 | 536 | 30 | 0 |
| 2020 | 10 | 3795 | 2263 | 711 | 572 | 178 | 110 | 27 | 1369852 | 699830 | 68242 | 27213 | 3694 | 1736 | 415 |
| 2021 | 10 | 774 | 1679 | 703 | 272 | 140 | 24 | 11 | 29371 | 129127 | 23776 | 3259 | 1173 | 30 | 10 |

Table 35.11. Continued.

| Comb-WCGFS-Q4 - Combined Scottish and Irish Groundfish Survey - numbers-at-age per hour |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Index |  |  |  |  |  |  |  |  |  |
| Year |  | Age |  |  |  |  |  |  |  |
|  | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1996 | 1 | 7924.6 | 449.5 | 99.5 | 30.0 | 7.9 | 3.1 | 0.5 | 0.0 |
| 1997 | 1 | 1848.2 | 701.9 | 155.3 | 33.5 | 10.3 | 2.3 | 0.4 | 0.1 |
| 1998 | 1 | 551.3 | 617.9 | 198.2 | 20.7 | 9.0 | 2.1 | 0.4 | 0.7 |
| 1999 | 1 | 2060.8 | 218.5 | 97.1 | 14.5 | 4.5 | 2.1 | 0.1 | 0.3 |
| 2000 | 1 | 2527.0 | 1035.4 | 161.4 | 25.0 | 1.4 | 1.2 | 0.2 | 0.4 |
| 2001 | 1 | 258.5 | 394.9 | 398.2 | 30.4 | 4.7 | 2.5 | 0.3 | 0.1 |
| 2002 | 1 | 2018.7 | 273.4 | 110.4 | 57.2 | 4.1 | 1.2 | 0.5 | 0.2 |
| 2003 | 1 | 558.1 | 485.5 | 74.0 | 32.9 | 10.3 | 2.1 | 0.4 | 0.2 |
| 2004 | 1 | 210.2 | 198.4 | 74.5 | 7.6 | 4.6 | 3.1 | 0.2 | 0.2 |
| 2005 | 1 | 166.8 | 68.2 | 43.3 | 10.5 | 1.0 | 0.4 | 0.1 | 0.1 |
| 2006 | 1 | 128.0 | 50.5 | 27.4 | 11.9 | 3.2 | 0.7 | 0.1 | 0.0 |
| 2007 | 1 | 111.7 | 59.9 | 39.2 | 11.3 | 4.5 | 2.6 | 0.3 | 0.1 |
| 2008 | 1 | 18.6 | 41.5 | 19.1 | 15.9 | 4.1 | 3.0 | 0.5 | 0.1 |
| 2009 | 1 | 1492.7 | 23.2 | 16.5 | 5.1 | 2.6 | 0.8 | 0.4 | 0.4 |
| 2010 | 1 | 76.0 | 317.9 | 35.2 | 8.8 | 2.5 | 1.1 | 0.2 | 0.3 |
| 2011 | 1 | 428.5 | 33.4 | 157.4 | 20.4 | 7.1 | 2.4 | 1.4 | 0.4 |
| 2012 | 1 | 89.3 | 241.6 | 50.9 | 69.7 | 16.6 | 2.4 | 0.7 | 0.3 |
| 2013 | 1 | 2140.4 | 38.3 | 90.5 | 27.9 | 29.7 | 4.9 | 0.6 | 0.1 |
| 2014 | 1 | 7693.9 | 243.8 | 46.2 | 38.3 | 10.2 | 8.1 | 2.0 | 0.2 |
| 2015 | 1 | 804.2 | 781.8 | 119.1 | 27.9 | 18.7 | 4.4 | 2.8 | 0.2 |
| 2016 | 1 | 542.8 | 310.0 | 288.6 | 39.8 | 8.7 | 10.5 | 1.6 | 2.2 |
| 2017 | 1 | 579.2 | 162.5 | 112.7 | 118.6 | 18.6 | 2.6 | 1.3 | 0.7 |
| 2018 | 1 | 2856.0 | 134.1 | 94.5 | 37.6 | 20.1 | 2.9 | 0.1 | 0.4 |
| 2019 | 1 | 2833.4 | 440.5 | 46.8 | 22.4 | 9.4 | 3.5 | 0.6 | 0.1 |
| 2020 | 1 | 1063.6 | 426.8 | 167.1 | 23.9 | 12.0 | 4.0 | 2.1 | 0.2 |
| 2021 | 1 | 1521.8 | 129.7 | 157.5 | 36.1 | 6.8 | 4.1 | 2.1 | 0.2 |

Table 35.11. Continued.

| Comb-WCGFS-Q4 - Combined Scottish and Irish Groundfish Survey - numbers-at-age per hour |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Variance |  |  |  |  |  |  |  |  |  |
| Year | Effort | Age |  |  |  |  |  |  |  |
|  | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1996 | 1 | 13802654.4 | 28791.0 | 1511.8 | 116.4 | 7.7 | 1.2 | 0.0 | 0.0 |
| 1997 | 1 | 503549.8 | 52750.2 | 2250.0 | 88.0 | 7.5 | 0.3 | 0.0 | 0.0 |
| 1998 | 1 | 51926.4 | 41240.3 | 4136.9 | 36.9 | 5.9 | 0.4 | 0.0 | 0.0 |
| 1999 | 1 | 696436.9 | 5025.7 | 904.9 | 17.2 | 1.7 | 0.5 | 0.0 | 0.0 |
| 2000 | 1 | 710042.2 | 80593.8 | 2267.5 | 41.2 | 0.2 | 0.1 | 0.0 | 0.0 |
| 2001 | 1 | 9203.0 | 14074.1 | 11808.5 | 49.2 | 1.2 | 0.3 | 0.0 | 0.0 |
| 2002 | 1 | 474926.4 | 5227.6 | 718.7 | 183.1 | 0.8 | 0.1 | 0.0 | 0.0 |
| 2003 | 1 | 27631.5 | 12731.6 | 247.4 | 38.6 | 3.0 | 0.1 | 0.0 | 0.0 |
| 2004 | 1 | 5010.9 | 2090.7 | 242.5 | 2.3 | 0.8 | 0.3 | 0.0 | 0.0 |
| 2005 | 1 | 2757.1 | 251.2 | 78.5 | 3.8 | 0.1 | 0.0 | 0.0 | 0.0 |
| 2006 | 1 | 2761.4 | 135.1 | 30.6 | 5.0 | 0.4 | 0.0 | 0.0 | 0.0 |
| 2007 | 1 | 1053.9 | 183.5 | 63.9 | 4.5 | 0.7 | 0.2 | 0.0 | 0.0 |
| 2008 | 1 | 75.6 | 98.1 | 18.1 | 11.5 | 0.6 | 0.3 | 0.0 | 0.0 |
| 2009 | 1 | 186968.9 | 32.7 | 12.7 | 1.2 | 0.3 | 0.0 | 0.0 | 0.0 |
| 2010 | 1 | 1318.7 | 10785.8 | 118.5 | 6.6 | 0.9 | 0.1 | 0.0 | 0.0 |
| 2011 | 1 | 22015.9 | 64.1 | 1135.1 | 14.5 | 1.6 | 0.2 | 0.1 | 0.0 |
| 2012 | 1 | 1072.7 | 3207.9 | 120.1 | 156.4 | 7.8 | 0.2 | 0.0 | 0.0 |
| 2013 | 1 | 547663.6 | 105.2 | 466.5 | 34.0 | 31.0 | 0.9 | 0.0 | 0.0 |
| 2014 | 1 | 5504945.1 | 2751.6 | 88.9 | 52.1 | 3.3 | 1.7 | 0.1 | 0.0 |
| 2015 | 1 | 45158.0 | 29135.7 | 580.9 | 31.2 | 13.6 | 0.6 | 0.2 | 0.0 |
| 2016 | 1 | 27390.7 | 4501.4 | 3657.3 | 57.9 | 2.7 | 3.1 | 0.1 | 0.1 |
| 2017 | 1 | 39781.0 | 1500.3 | 643.5 | 534.3 | 12.0 | 0.2 | 0.1 | 0.0 |
| 2018 | 1 | 666403.1 | 934.4 | 400.3 | 58.6 | 14.0 | 0.3 | 0.0 | 0.0 |
| 2019 | 1 | 655483.7 | 9093.9 | 88.3 | 19.8 | 3.3 | 0.4 | 0.0 | 0.0 |
| 2020 | 1 | 158671.1 | 9975.1 | 1333.3 | 22.8 | 5.4 | 0.5 | 0.1 | 0.0 |
| 2021 | 1 | 190708.3 | 784.4 | 1033.5 | 47.0 | 1.9 | 0.7 | 0.1 | 0.0 |

Table 35.12. Whiting in Division 6.a. SAM configuration settings for assessment of 6.a whiting agreed at WKNSEA 2021.

| Model Setting | Setting name | Configuration \& details |
| :---: | :---: | :---: |
| Minimum age in model | \$minage | 0 |
| Maximum age in model | \$maxAge | 7 |
| Maximum age plus group | \$maxAgePlusGroup | Maximum age plus group applies to both the commercial catch data and modelled Q4 survey index |
| Coupling of the fishing mortality states processes | \$keyLogFsta | Uncoupled across all age classes |
| Correlation of fishing mortality across ages | \$corFlag | $A R(1)$ first order autoregressive |
| Coupling of the survey catchability parameters | \$keyLogFpar | WCIBTS.Q1: ages 1 to 4 uncoupled; ages 5 and 6 coupled |
|  |  | SCO.Q1: ages 1 to 4 uncoupled; ages 5 and 6 coupled |
|  |  | SWC.Q4: ages 0 to 5 uncoupled; ages 6 and 7+ coupled |
| Density dependent catchability power parameters | \$keyQpow | $\mathrm{n} / \mathrm{a}$ |
| Coupling of process variance parameters for $\log (\mathrm{F})$ process | \$keyVarF | Coupled across all age classes |
| Coupling of the recruitment and survival process variance parameters | \$keyVarLogN | Age 0 uncoupled; ages 1 to 7+ coupled |
| Coupling of the variance parameters for the observations | \$keyVarObs | Catch: age 0 uncoupled; ages 1 to 6 coupled; age 7+ uncoupled |
|  |  | WCIBTS.Q1: ages 1 to 6 coupled |
|  |  | SCO.Q1: ages 1 to 6 coupled |
|  |  | SWC.Q4: age 0 uncoupled; ages 1 to 4 coupled; ages 5 to 7+ coupled |
| Covariance structure for each fleet | \$obsCorStruct | Catch: Independent ("ID") |
|  |  | WCIBTS.Q1: "ID" |
|  |  | SCO.Q1: "ID" |
|  |  | SWC.Q4: first order autoregressive ("AR1") |
| Coupling of correlation parameters for fleet covariance | \$keyCorObs | SWC.Q4: ages 0 and 1 coupled; ages 1 to 6 coupled; ages 6 and 7+ coupled |
| Stock recruitment code | \$stockRecruitmentModelCode | 0; Plain random walk |
| Number of years where catch scaling is applied | \$noScaledYears | 0 |
| Years where catch is scaled | \$keyScaledYears | $\mathrm{n} / \mathrm{a}$ |
| Matrix specifying the couplings of scale parameters | \$keyParScaledYA | $\mathrm{n} / \mathrm{a}$ |
| Lowest and highest ages included in $\bar{F}$ | \$fbarRange | 1,3 |


| Model Setting | Setting name | Configuration \& details |
| :---: | :---: | :---: |
| Biomass survey configuration | \$keyBiomassTreat | $\mathrm{n} / \mathrm{a}$ |
| Observational likelihood | \$obsLikelihoodFlag | Catch: "LN" |
|  |  | WCIBTS.Q1: "LN" |
|  |  | SCO.Q1: "LN" |
|  |  | SWC.Q4: "LN" |
| Observation weighting configuration | \$fixVarToWeight | 0 |
| Fraction of $t(3)$ distribution used in logF increment distribution | \$fracMixF | 0 |
| Fraction of $t(3)$ distribution used in $\log N$ increment distribution | \$fracMixN | 0 |
| Fraction of $\mathrm{t}(3)$ distribution used in distribution of fleets | \$fracMixObs | Catch: 0 |
|  |  | WCIBTS.Q1: 0 |
|  |  | SCO.Q1: 0 |
|  |  | SWC.Q4: 0 |
| Break years between which recruitment is constant | \$constRecBreaks | $\mathrm{n} / \mathrm{a}$ |
| Coupling of parameters used in a prediction-variance link for observations | \$predVarObsLink | $\mathrm{n} / \mathrm{a}$ |

Table 35.13. Whiting in Division 6.a. Parameter estimates from the updated SAM assessment model.

| Parameter name | par | sd(par) | exp(par) | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: |
| logFpar_0 | -5.904 | 0.157 | 0.003 | 0.002 | 0.004 |
| logFpar_1 | -5.937 | 0.158 | 0.003 | 0.002 | 0.004 |
| logFpar_2 | -6.091 | 0.161 | 0.002 | 0.002 | 0.003 |
| logFpar_3 | -6.199 | 0.169 | 0.002 | 0.001 | 0.003 |
| logFpar_4 | -6.834 | 0.186 | 0.001 | 0.001 | 0.002 |
| logFpar_5 | -5.864 | 0.239 | 0.003 | 0.002 | 0.005 |
| logFpar_6 | -5.677 | 0.253 | 0.003 | 0.002 | 0.006 |
| logFpar_7 | -5.463 | 0.25 | 0.004 | 0.003 | 0.007 |
| logFpar_8 | -5.672 | 0.26 | 0.003 | 0.002 | 0.006 |
| logFpar_9 | -6.289 | 0.259 | 0.002 | 0.001 | 0.003 |
| logFpar_10 | -4.782 | 0.222 | 0.008 | 0.005 | 0.013 |
| logFpar_11 | -5.152 | 0.183 | 0.006 | 0.004 | 0.008 |
| logFpar_12 | -5.142 | 0.182 | 0.006 | 0.004 | 0.008 |
| logFpar_13 | -5.535 | 0.187 | 0.004 | 0.003 | 0.006 |
| logFpar_14 | -5.935 | 0.206 | 0.003 | 0.002 | 0.004 |
| logFpar_15 | -6.221 | 0.253 | 0.002 | 0.001 | 0.003 |
| logFpar_16 | -7.026 | 0.302 | 0.001 | 0 | 0.002 |
| logSdLogFsta_0 | -1.025 | 0.141 | 0.359 | 0.271 | 0.475 |
| $\operatorname{logSdLogN}$ | -0.518 | 0.165 | 0.596 | 0.428 | 0.828 |
| logSdLogN_1 | -1.776 | 0.257 | 0.169 | 0.101 | 0.283 |
| logSdLogObs_0 | -0.15 | 0.205 | 0.861 | 0.571 | 1.297 |
| logSdLogObs_1 | -1.033 | 0.078 | 0.356 | 0.304 | 0.416 |
| logSdLogObs_2 | -0.579 | 0.165 | 0.56 | 0.403 | 0.78 |
| logSdLogObs_3 | -0.332 | 0.066 | 0.717 | 0.629 | 0.818 |
| logSdLogObs_4 | 0.745 | 0.099 | 2.107 | 1.729 | 2.567 |
| logSdLogObs_5 | 0.974 | 0.15 | 2.648 | 1.96 | 3.576 |
| logSdLogObs_6 | 0.797 | 0.133 | 2.218 | 1.701 | 2.893 |
| logSdLogObs_7 | 1.172 | 0.101 | 3.228 | 2.635 | 3.954 |
| transfIRARdist_0 | 3.39 | 1212.843 | 29.662 | 0 | Inf |
| transfIRARdist_1 | -0.947 | 0.301 | 0.388 | 0.212 | 0.709 |
| transfIRARdist_2 | 1.595 | 2.195 | 4.929 | 0.061 | 397.639 |
| itrans_rho_0 | 1.565 | 0.195 | 4.785 | 3.24 | 7.067 |

* The relatively large standard deviation (and associated uncertainty) around the estimate of transfirARdist_0, the coupled AR1 parameter for ages 0 and 1, indicates a weak to non-existent level of autocorrelation between age groups 0 and 1.

Table 35.14. Whiting in Division 6.a. SAM estimated population numbers-at-age (thousands).

| Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 792418 | 193653 | 467016 | 80851 | 22098 | 6848 | 2073 | 1698 |
| 1982 | 816452 | 270104 | 70817 | 208307 | 35534 | 9460 | 2994 | 1846 |
| 1983 | 1068364 | 270655 | 79582 | 31142 | 87623 | 14366 | 3776 | 2148 |
| 1984 | 876855 | 370063 | 79473 | 26908 | 11215 | 30726 | 4537 | 2227 |
| 1985 | 744978 | 289359 | 111446 | 26553 | 7582 | 3245 | 8353 | 2064 |
| 1986 | 1338213 | 231131 | 95963 | 36063 | 6987 | 1626 | 676 | 2278 |
| 1987 | 502134 | 506331 | 87467 | 36869 | 12398 | 2112 | 474 | 1117 |
| 1988 | 752010 | 145880 | 157505 | 31655 | 11132 | 3334 | 501 | 540 |
| 1989 | 651153 | 251470 | 38874 | 45431 | 9925 | 2593 | 688 | 291 |
| 1990 | 852259 | 202756 | 88564 | 11157 | 12736 | 2876 | 572 | 261 |
| 1991 | 1131396 | 279056 | 65336 | 39039 | 4010 | 4238 | 924 | 305 |
| 1992 | 767342 | 391116 | 97846 | 24148 | 16014 | 1436 | 1343 | 469 |
| 1993 | 782199 | 243165 | 134805 | 35727 | 8756 | 6116 | 501 | 715 |
| 1994 | 755541 | 259288 | 88621 | 51249 | 13297 | 3096 | 2157 | 527 |
| 1995 | 682976 | 251299 | 101668 | 37070 | 18635 | 4711 | 1090 | 1077 |
| 1996 | 640194 | 222335 | 91449 | 37763 | 13506 | 5846 | 1521 | 873 |
| 1997 | 654177 | 198282 | 66622 | 32589 | 11676 | 3721 | 1430 | 840 |
| 1998 | 398209 | 218410 | 62625 | 20507 | 10947 | 3214 | 1015 | 912 |
| 1999 | 865524 | 113711 | 64480 | 15663 | 5400 | 2823 | 661 | 661 |
| 2000 | 326837 | 311800 | 37386 | 16690 | 3439 | 1235 | 620 | 454 |
| 2001 | 204756 | 92571 | 87084 | 11463 | 3589 | 833 | 262 | 335 |
| 2002 | 274808 | 62080 | 33075 | 26024 | 3676 | 1081 | 231 | 254 |
| 2003 | 214958 | 87095 | 16699 | 12139 | 7615 | 1335 | 417 | 219 |
| 2004 | 118957 | 71387 | 27011 | 5176 | 4062 | 2768 | 488 | 332 |
| 2005 | 109455 | 35075 | 20933 | 8354 | 1540 | 1282 | 957 | 400 |
| 2006 | 68001 | 36103 | 12326 | 8171 | 3334 | 687 | 626 | 749 |
| 2007 | 60957 | 20606 | 13804 | 5516 | 3434 | 1584 | 342 | 787 |
| 2008 | 103708 | 19403 | 7553 | 6661 | 2560 | 1739 | 746 | 638 |
| 2009 | 303805 | 35430 | 8788 | 3805 | 3386 | 1285 | 914 | 776 |
| 2010 | 95360 | 114222 | 14490 | 4472 | 1884 | 1857 | 661 | 1001 |


| Year | Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 200140 | 26096 | 42496 | 7282 | 2331 | 924 | 1047 | $\mathbf{7 +}$ |
| 2012 | 108065 | 70825 | 11253 | 19228 | 4603 | 1137 | 530 | 1045 |
| 2013 | 243043 | 33420 | 25621 | 6616 | 10084 | 2338 | 659 | 911 |
| 2014 | 447703 | 79396 | 9898 | 10977 | 3909 | 4805 | 1347 | 984 |
| 2015 | 307589 | 150437 | 26402 | 6299 | 6507 | 2227 | 3000 | 1439 |
| 2016 | 261930 | 97065 | 64831 | 12813 | 3486 | 3699 | 1300 | 2621 |
| 2017 | 194799 | 90460 | 37284 | 33126 | 6894 | 2038 | 2024 | 2046 |
| 2018 | 396956 | 57931 | 36790 | 19429 | 16009 | 3709 | 1060 | 2101 |
| 2020 | 433379 | 140808 | 22105 | 18357 | 10308 | 7759 | 2098 | 1681 |
| 2021 | 273676 | 60682 | 57372 | 23059 | 6912 | 6179 | 3266 | 3921 |

Table 35.15. Whiting in Division 6.a. SAM estimates for F-at-age

| Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1981 | 0.022 | 0.06 | 0.128 | 0.221 | 0.279 | 0.318 | 0.343 | 0.178 |
| 1982 | 0.033 | 0.093 | 0.194 | 0.315 | 0.38 | 0.424 | 0.454 | 0.232 |
| 1983 | 0.046 | 0.135 | 0.291 | 0.471 | 0.565 | 0.635 | 0.675 | 0.335 |
| 1984 | 0.054 | 0.161 | 0.364 | 0.602 | 0.74 | 0.835 | 0.901 | 0.437 |
| 1985 | 0.058 | 0.174 | 0.421 | 0.714 | 0.932 | 1.068 | 1.171 | 0.548 |
| 1986 | 0.045 | 0.132 | 0.312 | 0.495 | 0.643 | 0.743 | 0.842 | 0.393 |
| 1987 | 0.069 | 0.211 | 0.455 | 0.647 | 0.811 | 0.938 | 1.05 | 0.468 |
| 1988 | 0.083 | 0.259 | 0.537 | 0.719 | 0.908 | 1.074 | 1.177 | 0.512 |
| 1989 | 0.075 | 0.232 | 0.483 | 0.651 | 0.801 | 0.963 | 1.032 | 0.439 |
| 1990 | 0.055 | 0.167 | 0.333 | 0.445 | 0.544 | 0.641 | 0.663 | 0.285 |
| 1991 | 0.055 | 0.166 | 0.337 | 0.452 | 0.54 | 0.613 | 0.611 | 0.263 |
| 1992 | 0.053 | 0.16 | 0.324 | 0.437 | 0.515 | 0.566 | 0.554 | 0.241 |
| 1993 | 0.044 | 0.13 | 0.272 | 0.399 | 0.505 | 0.57 | 0.554 | 0.237 |
| 1994 | 0.04 | 0.117 | 0.245 | 0.37 | 0.478 | 0.536 | 0.518 | 0.217 |
| 1995 | 0.051 | 0.155 | 0.314 | 0.457 | 0.575 | 0.635 | 0.592 | 0.241 |
| 1996 | 0.07 | 0.219 | 0.424 | 0.609 | 0.748 | 0.82 | 0.716 | 0.274 |
| 1997 | 0.071 | 0.221 | 0.413 | 0.586 | 0.688 | 0.746 | 0.627 | 0.23 |
| 1998 | 0.097 | 0.311 | 0.567 | 0.798 | 0.901 | 0.992 | 0.839 | 0.289 |
| 1999 | 0.094 | 0.304 | 0.559 | 0.827 | 0.954 | 1.022 | 0.88 | 0.282 |
| 2000 | 0.11 | 0.359 | 0.58 | 0.849 | 1.004 | 0.995 | 0.836 | 0.252 |
| 2001 | 0.08 | 0.255 | 0.408 | 0.584 | 0.71 | 0.67 | 0.521 | 0.15 |
| 2002 | 0.082 | 0.263 | 0.371 | 0.497 | 0.558 | 0.483 | 0.333 | 0.096 |
| 2003 | 0.079 | 0.251 | 0.324 | 0.439 | 0.484 | 0.401 | 0.264 | 0.076 |
| 2004 | 0.101 | 0.323 | 0.364 | 0.498 | 0.53 | 0.419 | 0.272 | 0.083 |
| 2005 | 0.058 | 0.177 | 0.167 | 0.226 | 0.243 | 0.193 | 0.13 | 0.043 |
| 2006 | 0.052 | 0.161 | 0.133 | 0.189 | 0.224 | 0.192 | 0.139 | 0.048 |
| 2007 | 0.041 | 0.131 | 0.101 | 0.148 | 0.19 | 0.17 | 0.125 | 0.046 |
| 2008 | 0.034 | 0.111 | 0.074 | 0.109 | 0.147 | 0.135 | 0.098 | 0.037 |
| 2009 | 0.049 | 0.171 | 0.089 | 0.118 | 0.15 | 0.131 | 0.09 | 0.034 |
| 2010 | 0.064 | 0.229 | 0.103 | 0.123 | 0.147 | 0.131 | 0.089 | 0.034 |


| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 2011 | 0.039 | 0.14 | 0.051 | 0.056 | 0.064 | 0.057 | 0.038 | 0.015 |
| 2012 | 0.048 | 0.186 | 0.059 | 0.063 | 0.071 | 0.065 | 0.041 | 0.015 |
| 2013 | 0.058 | 0.244 | 0.065 | 0.065 | 0.07 | 0.063 | 0.04 | 0.015 |
| 2014 | 0.03 | 0.122 | 0.032 | 0.035 | 0.04 | 0.041 | 0.027 | 0.01 |
| 2015 | 0.034 | 0.138 | 0.04 | 0.048 | 0.053 | 0.056 | 0.034 | 0.013 |
| 2016 | 0.027 | 0.112 | 0.033 | 0.039 | 0.042 | 0.046 | 0.025 | 0.009 |
| 2017 | 0.027 | 0.119 | 0.034 | 0.043 | 0.046 | 0.052 | 0.026 | 0.008 |
| 2018 | 0.021 | 0.094 | 0.027 | 0.035 | 0.038 | 0.039 | 0.018 | 0.006 |
| 2019 | 0.026 | 0.121 | 0.034 | 0.046 | 0.051 | 0.051 | 0.021 | 0.006 |
| 2020 | 0.026 | 0.122 | 0.03 | 0.04 | 0.045 | 0.044 | 0.017 | 0.005 |
| 2021 | 0.027 | 0.13 | 0.033 | 0.048 | 0.061 | 0.064 | 0.025 | 0.007 |

Table 35.16. Whiting in Division 6.a. Assessment summary with weights in tonnes and recruitment in thousands. 'High' and 'Low' refer to $95 \%$ confidence intervals.

| Year | Recruitment age 0 |  |  | SSB |  |  | Landings* | Discards* | Fishing mortality ages 1-3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Value | High | Low | Value | High |  |  | Low | Value | High |
| 1981 | 459996 | 792418 | 1365066 | 77786 | 108518 | 151393 | 12194 | 2132 | 0.091 | 0.136 | 0.204 |
| 1982 | 498421 | 816452 | 1337413 | 68579 | 91467 | 121993 | 13880 | 5485 | 0.142 | 0.201 | 0.283 |
| 1983 | 657345 | 1068364 | 1736379 | 52550 | 66838 | 85011 | 15962 | 6294 | 0.217 | 0.299 | 0.412 |
| 1984 | 539881 | 876855 | 1424156 | 37862 | 46702 | 57605 | 16459 | 4017 | 0.275 | 0.375 | 0.512 |
| 1985 | 451474 | 744978 | 1229290 | 30455 | 37889 | 47137 | 12879 | 4840 | 0.324 | 0.436 | 0.588 |
| 1986 | 807581 | 1338213 | 2217501 | 25504 | 32190 | 40629 | 8458 | 2669 | 0.228 | 0.313 | 0.428 |
| 1987 | 305224 | 502134 | 826076 | 28783 | 35679 | 44227 | 11542 | 11918 | 0.328 | 0.438 | 0.583 |
| 1988 | 446791 | 752010 | 1265733 | 31030 | 39873 | 51236 | 11349 | 8132 | 0.381 | 0.505 | 0.67 |
| 1989 | 402020 | 651153 | 1054675 | 20737 | 26023 | 32657 | 7523 | 5876 | 0.335 | 0.455 | 0.618 |
| 1990 | 529516 | 852259 | 1371717 | 19138 | 24796 | 32126 | 5642 | 4530 | 0.229 | 0.315 | 0.434 |
| 1991 | 700276 | 1131396 | 1827931 | 21377 | 27048 | 34224 | 6657 | 4883 | 0.233 | 0.318 | 0.434 |
| 1992 | 476492 | 767342 | 1235727 | 25955 | 32522 | 40749 | 6004 | 9249 | 0.224 | 0.307 | 0.421 |
| 1993 | 487073 | 782199 | 1256149 | 29801 | 37808 | 47966 | 6871 | 4759 | 0.193 | 0.267 | 0.368 |
| 1994 | 474415 | 755541 | 1203255 | 28991 | 36153 | 45084 | 5900 | 3455 | 0.176 | 0.244 | 0.338 |
| 1995 | 431387 | 682976 | 1081293 | 29491 | 36487 | 45144 | 6078 | 5771 | 0.227 | 0.309 | 0.42 |
| 1996 | 403631 | 640194 | 1015406 | 27181 | 33447 | 41156 | 7158 | 7940 | 0.31 | 0.417 | 0.561 |
| 1997 | 419932 | 654177 | 1019088 | 21653 | 26670 | 32849 | 6291 | 5251 | 0.3 | 0.407 | 0.551 |
| 1998 | 251844 | 398209 | 629639 | 18297 | 22542 | 27771 | 4628 | 9216 | 0.417 | 0.559 | 0.747 |
| 1999 | 547489 | 865524 | 1368304 | 14208 | 17817 | 22344 | 4613 | 3975 | 0.418 | 0.563 | 0.759 |
| 2000 | 207196 | 326837 | 515561 | 12326 | 15340 | 19091 | 3011 | 13285 | 0.445 | 0.596 | 0.798 |
| 2001 | 123439 | 204756 | 339642 | 13389 | 17587 | 23100 | 2439 | 4263 | 0.298 | 0.416 | 0.579 |
| 2002 | 172978 | 274808 | 436584 | 10429 | 13606 | 17752 | 1768 | 2851 | 0.26 | 0.377 | 0.548 |


| Year | Recruitment age 0 |  |  | SSB |  |  | Landings* | Discards* | Fishing mortality ages 1-3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | Value | High | Low | Value | High |  |  | Low | Value | High |
| 2003 | 138692 | 214958 | 333162 | 7411 | 9806 | 12976 | 1331 | 1987 | 0.224 | 0.338 | 0.511 |
| 2004 | 75301 | 118957 | 187922 | 6531 | 8857 | 12011 | 799 | 2889 | 0.252 | 0.395 | 0.621 |
| 2005 | 69310 | 109455 | 172852 | 5175 | 7333 | 10389 | 335 | 971 | 0.121 | 0.19 | 0.299 |
| 2006 | 42299 | 68001 | 109321 | 4684 | 6631 | 9386 | 378 | 748 | 0.105 | 0.161 | 0.248 |
| 2007 | 37923 | 60957 | 97979 | 4489 | 6399 | 9120 | 481 | 367 | 0.082 | 0.127 | 0.195 |
| 2008 | 63495 | 103708 | 169388 | 4067 | 5848 | 8409 | 441 | 156 | 0.063 | 0.098 | 0.152 |
| 2009 | 182966 | 303805 | 504451 | 4093 | 5870 | 8419 | 480 | 826 | 0.082 | 0.126 | 0.193 |
| 2010 | 56780 | 95360 | 160152 | 5395 | 7753 | 11143 | 338 | 1094 | 0.097 | 0.152 | 0.237 |
| 2011 | 121253 | 200140 | 330352 | 7755 | 11738 | 17766 | 229 | 631 | 0.051 | 0.082 | 0.134 |
| 2012 | 63597 | 108065 | 183626 | 8077 | 12242 | 18557 | 304 | 772 | 0.063 | 0.103 | 0.169 |
| 2013 | 143857 | 243043 | 410615 | 8647 | 13389 | 20730 | 216 | 1225 | 0.073 | 0.125 | 0.214 |
| 2014 | 266623 | 447703 | 751764 | 7749 | 11995 | 18569 | 181 | 748 | 0.037 | 0.063 | 0.106 |
| 2015 | 182614 | 307589 | 518092 | 9838 | 15192 | 23458 | 223 | 1457 | 0.045 | 0.075 | 0.125 |
| 2016 | 156219 | 261930 | 439172 | 14014 | 22004 | 34551 | 226 | 1038 | 0.036 | 0.061 | 0.104 |
| 2017 | 111597 | 194799 | 340036 | 15845 | 24777 | 38744 | 178 | 1326 | 0.039 | 0.065 | 0.11 |
| 2018 | 232604 | 396956 | 677437 | 15732 | 24363 | 37728 | 190 | 648 | 0.031 | 0.052 | 0.088 |
| 2019 | 248326 | 433379 | 756332 | 14861 | 22780 | 34917 | 502 | 925 | 0.04 | 0.067 | 0.113 |
| 2020 | 102183 | 199449 | 389299 | 16902 | 26085 | 40258 | 544 | 826 | 0.036 | 0.064 | 0.113 |
| 2021 | 93253 | 273676 | 803174 | 18547 | 29141 | 45786 | 873 | 262*** | 0.036 | 0.07 | 0.139 |
| 2022 | 108065 | 273676** | 447703 | 17660 | 29167 | 47163 |  |  |  |  |  |

* Calculated using Sum of Products from the catch numbers-at-age and mean weights-at-age. Pre-2003 Discards are estimated for ages 1+ only.


## ** Median resampled recruitment (2012-2021).

*** Underestimate due to reduced discard sampling from the Nephrops fleet.

Table 35.17. Whiting in Division 6.a. Annual catch scenarios. All weights are in tonnes.

| Basis | $\begin{aligned} & \text { Total } \\ & \text { catch } \\ & (2023) \end{aligned}$ | ```Projected landings * (2023)``` | ```Projected discards ** (2023)``` | $\begin{gathered} F_{\text {total }} \\ (2023) \end{gathered}$ | $\mathrm{F}_{\text {projected land- }}$ <br> ings (2023) | $\mathrm{F}_{\text {projected dis- }}$ cards (2023) | $\begin{gathered} \text { SSB } \\ (2024) \end{gathered}$ | $\begin{gathered} \text { \% SSB } \\ \text { change }{ }^{* * *} \end{gathered}$ | \% Advice change^ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY approach $=$ $\mathrm{F}_{\mathrm{MSY}}$ | 4155 | 2081 | 2074 | 0.21 | 0.043 | 0.167 | 25692 | -10.6 | 1 |
| $F=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 30665 | 6.7 | -100 |
| FMSY lower | 3472 | 1735 | 1737 | 0.173 | 0.036 | 0.137 | 26463 | -7.9 | -15.6 |
| $\mathrm{F}_{\text {MSY upper }}$ | 4155 | 2081 | 2074 | 0.21 | 0.043 | 0.167 | 25692 | -10.6 | 1 |
| $\mathrm{F}_{\mathrm{pa}}$ | 4155 | 2081 | 2074 | 0.21 | 0.043 | 0.167 | 25692 | -10.6 | 1 |
| $\mathrm{F}_{\text {lim }}$ | 5907 | 2978 | 2929 | 0.31 | 0.064 | 0.25 | 23765 | -17.3 | 44 |
| SSB (2023) $=$ B $_{\text {lim }}$ | 12447 | 6432 | 6015 | 0.77 | 0.158 | 0.61 | 17286 | -40 | 200 |
| SSB (2023) $=\mathrm{B}_{\mathrm{pa}}$ | 4241 | 2124 | 2117 | 0.22 | 0.044 | 0.171 | 25597 | -10.90 | 3.1 |
| $\begin{aligned} & \text { SSB (2023) }=\text { MSY } \\ & \text { Btrigger }^{\text {the }} \end{aligned}$ | 4241 | 2124 | 2117 | 0.22 | 0.044 | 0.171 | 25597 | -10.90 | 3.1 |
| $\begin{aligned} & \text { SSB (2024) = SSB } \\ & (2023) \end{aligned}$ | 1547 | 766 | 781 | 0.074 | 0.0153 | 0.059 | 28727 | 0 | -62 |
| $F=F_{2022}$ | 1468 | 728 | 740 | 0.070 | 0.0145 | 0.055 | 28825 | 0.34 | -64 |

* Marketable landings, assuming recent discard rate.
** Including BMS landings (EU stocks), assuming recent discard rate.
*** SSB 2024 relative to SSB 2023.



Figure 35.1. Whiting in Division 6.a. Landings and discards (in thousand tonnes) as officially reported to ICES (upper panel) and discards (as \% of catch, lower panel). Pre-2003 discards are estimated for ages 1+ only; from 2003 onwards, they are estimated for all ages.
whg.27.6a LandWt

whg.27.6a DisWt


Figure 35.2. Whiting in Division 6.a. Landings (upper panel) and discards (all ages, lower panel) by métier (kg) in 2019 as entered into InterCatch.
whg.27.6a CatchWt


Figure 35.3. Whiting in Division 6.a. Landings (sampled and unsampled, in grey), sampled discards (in black) and raised unsampled discards (in red) after allocations within InterCatch.

Total Catch Numbers At Age


Figure 35.4. Whiting in Division 6.a. Catch numbers-at-age by sampled and unsampled landings and sampled and raised (unsampled) discards, after allocations within InterCatch.


Figure 35.5. Whiting in Division 6.a. Catch numbers-at-age by year.


Figure 35.6. Whiting in Division 6.a. Mean weight-at-age in the landings, discards, catch and stock.


Figure 35.7. Whiting in Division 6.a. Catch numbers-at-age by sampled and un-sampled landings and sampled and raised (unsampled) discards, after allocations within InterCatch, in 2021 and in recent years 2017-2020.


Figure 35.8. Whiting in Division 6.a. Combined Q1 and Q4 survey weights-at-age time-series for 6a whiting, together with catch weights-at-age time-series. Only Q4 surveys contain data for the zero age class. The smoothed estimates were used for M calculation.


Figure 35.9. Whiting in Division 6.a. Time-series of natural mortality-at-age estimated with Lorenzen's (1996) model. The thick black line shows the natural mortality obtained with the smoothed weights-at-age with the corresponding 95\% confidence interval shown in grey. The thin black line shows the natural mortality obtained with unsmoothed weights-at-age, for comparison.

UK-SCOWCGFS-Q1 \& UK-SCOWCGFS-Q4: whiting


Figure 35.10. Whiting in Division 6.a. CPUE from the Scottish first quarter west coast groundfish survey (UK-SCOWCGFSQ1, in red) and the Scottish fourth quarter groundfish survey (UK-SCOWCGFS-Q4, in blue) in 2017-2021. Numbers are standardised to $\mathbf{3 0}$ minutes towing. The closed areas, the Windsock and West Shetland Shelf in the north, and the Clyde in the south, are shown as green polygons.


Figure 35.11. Whiting in Division 6.a. CPUE from the Scottish fourth quarter west coast groundfish survey (UK-SCOWCGFSQ4, only the southern part of the survey area, in blue) and the Irish fourth quarter groundfish survey (IGFS-WIBTS-Q4, in green) in 2017-2021. Numbers are standardised to $\mathbf{3 0}$ minutes towing. The Clyde closed area is shown as areen polygon.


Figure 35.12. Whiting in Division 6.a. The combined index derived from a delta-GAM model fit to data from the three Q4 surveys (black line) with $95 \%$ confidence limits (in grey). Indices are derived by summing model predictions on a spatial grid. The survey index calculated using the stratified mean method for ICES statistical rectangles as strata are shown as red points. The indices are mean-standardised.


Figure 35.13. Whiting in Division 6.a. Log abundance indices by year with a line for each cohort, for catch and the three survey series. The spawning year of each cohort is indicated at the start of each line. Note the age range 1-7+ in 1981-2002 and 0-7+ in 2003-2021 for the catch data.


Figure 35.14. Whiting in Division 6.a. Standardised proportions-at-age per year ("spay") for the three survey series. The positive values are shown in red, the negative values are shown in blue.




Figure 35.15. Whiting in Division 6.a. Within-survey correlations comparing index values at different ages for the same year classes for the three survey series. The straight line is a linear regression.


Figure 35.16. Whiting in Division 6.a. Summary of the SAM assessment model estimates (orange line) with 95\% confidence intervals (yellow polygon).


Figure 35.17. Whiting in Division 6.a. Comparison of the SAM assessment model estimates with observed log catch num-bers-at-age (first column of panels) and observed log survey indices-at-age for ScoGFS-WIBTS-Q1 (second column), Comb-WCGFS-Q4 (third column), and UK-SCOWCGFS-Q1 (fourth column).


Figure 35.18. Whiting in Division 6.a. Standardized one-observation-ahead residuals-at-age by fleet from the SAM assessment model: catch (top left), ScoGFS-WIBTS-Q1 (bottom left), UK-SCOWCGFS-Q1 (bottom right), and Comb-WCGFS-Q4 (top right).


Figure 35.19. Whiting in Division 6.a. Leave-one-out sensitivity analysis of the SAM assessment model.


Figure 35.20. Whiting in Division 6.a. Retrospective patterns for the SAM assessment model.


Figure 35.21. Whiting in Division 6.a. Sensitivity analysis investigating the effect of missing catch numbers-at-age zero and one value in the terminal year (solid lines), when compared to the full model retrospective fits (dotted lines).


Figure 35.22. Whiting in Division 6.a. The SAM assessment model Stock-Recruit relationship.


Figure 35.23. Whiting in Division 6.a. Comparison of 2022 SAM assessment estimates (solid lines) with revised 2021 SAM assessment estimates (dotted lines).


Figure 35.24. Whiting in Division 6.a. Results of SURBAR analysis (see legend on mean Z plot for details). SSB, TSB and recruitment are relative estimates.


Figure 35.25. Whiting in Division 6.a. Comparison of the Recruitment and SSB estimates by SAM and SURBAR (the run with three tuning series).


Figure 35.26. Whiting in Division 6.a. ICES Standard Graphs for the SAM assessment.


Figure 35.27. Whiting in Division 6.a. The SAM assessment model estimated F-at-age.


Figure 35.28. Whiting in Division 6.a. Proportion of catch discarded-at-age, from SAM landing fraction input file.


Figure 35.29. Whiting in Division 6.a. SAM forecast in the intermediate year followed by F $_{\text {MSY }}$ (the proposed advice) in subsequent years.


Figure 35.30. Whiting in Division 6.a. Contribution of recruitment years to projected 2022 catch and 2023 SSB under an $\mathrm{F}_{\text {MSY }}$ catch scenario.

## 36 Whiting (Merlangius merlangus) in Division 6.b (Rockall)

## Type of assessment in 2022

No assessment was performed in 2021.

## ICES advice applicable to 2022

In 2021, ICES provided multiyear advice:
"ICES advises that when the precautionary approach is applied, landings should be no more than 7 tonnes in each of the years 2022, 2023, 2024. ICES cannot quantify the corresponding total catches."
https://www.ices.dk/sites/pub/Publication\ Reports/Advice/2021/2021/whg.27.6b.pdf

## ICES advice applicable to 2021

In 2018, ICES provided multiyear advice:
"ICES advises that when the precautionary approach is applied, wanted catches should be no more than 9 tonnes in each of the years 2019, 2020, and 2021. ICES cannot quantify the corresponding total catches."
http://www.ices.dk/sites/pub/Publication\ Reports/Advice/2018/2018/whg.27.6b.pdf

### 36.1 General

## Stock description

There is an absence of information on whiting stock structure in this region and whiting caught at Rockall may potentially be part of the adjacent 6.a stock.

## Management applicable to 2021 and 2022

The TAC for whiting (in tonnes) is set for ICES subareas 6, 12 and 14 and EU and international waters of ICES Division 5b, for 2022-2020 is shown below.

TAC for 2022

(Annex IA to Regulation (EU) 2022/109).

## TAC for 2021

The agreed TAC was not available.

## TAC for 2020

| Species: | Whiting <br> Merlangius merlangus | Zone: | 6; Union and international waters of 5b; <br> international waters of 12 and 14 <br> (WHG/56-14) |
| :--- | :--- | :--- | :--- |
| Germany | $3\left(^{1}\right)$ | Analytical TAC |  |
| France | 57 | $\left({ }^{1}\right)$ | Article 8 of this Regulation applies |
| Ireland | 273 | $\left({ }^{1}\right)$ | Article 3 of Regulation (EC) No $847 / 96$ shall not apply |
| United Kingdom | 604 | $\left({ }^{1}\right)$ | Article 4 of Regulation (EC) No 847/96 shall not apply |
| Union | 937 | $\left({ }^{1}\right)$ |  |
| TAC | 937 | $\left({ }^{1}\right)$ |  |

$\left({ }^{1}\right)$ Exclusively for by-catches of whiting in fisheries for other species. No directed fisheries for whiting are permitted under this quota.
(Council Regulation (EU) 2020/123).

## Fishery in 2021

No specific information is available for 2021. Whiting at Rockall are taken as a bycatch in fisheries for other species such as haddock and anglerfish.

### 36.2 Data

Landings data for whiting in 27.6.b are shown by nation in Table 36.1 and Figure 36.1. Total officially reported landings were 17 t in 2021, of which 11 t were reported by the UK and 6 t by Ireland. In the past, official landings have shown very high interannual variation and it is not known whether these are a true reflection of removals.

Both landings and discards have been uploaded to InterCatch for 2021 (Figure 36.2). No information on the age composition in landings was available and some discard age compositions
(based on three sampled trips) were uploaded to InterCatch. All the landings and discards were from the Scottish TR1 fleet. The discard rate was $23 \%$. The data available in InterCatch are shown below.

| Country | Landings(tonnes) | Discards (tonnes) | Total (tonnes) |
| :--- | :---: | :---: | :---: |
| Ireland | 5.8 | 0.6 | 6.4 |
| UK (Scotland) | 11.1 | 4.4 | 15.5 |
| Grand total | 16.9 | 5.0 | 21.8 |

Survey catch rates of whiting at Rockall are extremely low (Table 36.2) and are therefore unlikely to provide a reliable index of abundance.

Catches of whiting (both survey and commercial) are too low to support the collection of the necessary information for an assessment of stock status.

### 36.3 Target category

In 2012, advice was provided using the DL approach for category 6; stocks with negligible landings stocks and stocks caught in minor amounts as bycatch with no indication of F in relation to reference points and no marked positive trends in stock indicators. WKLIFE has previously suggested a target category of 4 for this stock. Given the information in Section 36.2 regarding the potential unreliability of landings data and lack of sampled data, WGCSE considers that whiting in 27.6.b is likely to remain a category 6 stock.

### 36.4 Management considerations

Rockall whiting is managed under a TAC for the combined Divisions 6.a and 6.b and therefore cannot be effective in limiting catches in Rockall.

Table 36.1. Whiting in Division 27.6.b. Nominal landings (in tonnes) as officially reported to ICES.

| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| France | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Ireland | - | - | - | - | 32 | 10 | 4 | 23 | 3 | 1 | - | - | 10 | - | 2 | 3 | 3 | 104 |
| Norway | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Spain | - | - | - | - | - | - | - | - | - | - | + | - | - | - | - | - | - | - |
| UK (E, W \& NI) | 16 | 6 | 1 | 5 | 10 | 2 | 5 | 26 | 49 | 20 | - | - | - | - | - | - | - | - |
| UK (Scotland) | 18 | 482 | 459 | 283 | 86 | 68 | 53 | 36 | 65 | 23 | 44 | 58 | 4 | 7 | 11 | 1 | 1 | 1 |
| UK (all) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 34 | 488 | 460 | 288 | 128 | 80 | 62 | 85 | 117 | 44 | 44 | 58 | 14 | 7 | 13 | 4 | 4 | 105 |
| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020* | 2021* |  |  |  |
| Faroe Islands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |  |
| France | + | - | - | - | - | - | - | - | - | - | - | - | $+$ | $+$ | - |  |  |  |
| Ireland | 16 | 23 | 4 | 2 | 3 | - | + | 6 | 6 | 9 | 7 | 9 | 24 | 13 | 6 |  |  |  |
| Norway | - | - | - | - | - | - | - | - | - | 1 | - | + | - | - | - |  |  |  |
| Spain | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |  |
| UK (E, W \& NI) | - | - | - | - | - | - | - | - |  |  |  |  |  |  |  |  |  |  |
| UK (Scotland) | 1 | 8 | 12 | 16 | 6 | 1 | 3 | 23 |  |  |  |  |  |  |  |  |  |  |
| UK (all) |  |  |  |  |  |  |  |  | 46 | 22 | 32 | 34 | 65 | 25 | 11 |  |  |  |
| Total | 17 | 31 | 16 | 18 | 9 | 1 | 3 | 29 | 52 | 33 | 40 | 43 | 89 | 38 | 17 |  |  |  |

## * Preliminary

$+<0.5 \mathrm{t}$.

Table 36.2. Whiting in Division 27.6.b. Survey data made available to the WG: Scottish Q3 groundfish survey (UK-SCORocQ3). Catch rates are given as number per ten hours.

| Year | UK-SCORoc-Q3 - Scottish Groundfish Survey - numbers at age/10 h |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |  |
|  | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 2011 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 10 | 33.279 | 0 | 0.358 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 10 | 6.687 | 1.924 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 10 | 17.425 | 3.426 | 0.838 | 0.307 | 0 | 0 | 0 | 0 |
| 2015 | 10 | 8.853 | 0.559 | 0.559 | 0.55 | 0 | 0 | 0 | 0 |
| 2016 | 10 | 250.012 | 0.782 | 0 | 0.223 | 0.447 | 0 | 0 | 0 |
| 2017 | 10 | 23.147 | 10.84 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 10 | 0.531 | 0.754 | 0.894 | 0 | 0 | 0.307 | 0 | 0 |
| 2019 | 10 | 0.144 | 0.169 | 0.175 | 0 | 0 | 0.094 | 0 | 0 |
| 2020 | 10 | 9.388 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 10 | 26.112 | 0.902 | 0 | 0 | 0 | 0.223 | 0 | 0 |



Figure 36.1. Whiting in Division 27.6.b. Official landings of whiting in 27.6.b by nation.


Figure 36.2. Whiting in Division 27.6.b. Landings (left panel) and discards (right panel) by métier (kg) in 2021 as entered into InterCatch.

## 37 Whiting in 7.a (Irish Sea)

## 2020 Assessment and advice

WGCSE 2022 updated the assessment with 2021 data. The advice for this stock is biennial so advice does not change; however, a short-term forecast was run to update the data.

This stock was benchmarked in 2017 and the outcome was to upgrade the assessment from category 3 (trends based) to category 1 (analytical assessment and forecast). Data exploration was carried out in WKIrish 2 (ICES, 2017). A full analytical assessment procedure was developed during WKIrish 3 (ICES, 2017) using ASAP. Reference points were also estimated during WKIrish 3.

The advice for this stock was updated in October 2018 following a special request to ICES to update the advice based on the most recent discard estimates. Furthermore, in response to an EC request for advice on the removal of TACs for certain stocks ICES advised that removing the EU TAC for Whiting in ICES Division 7.a may generate a high risk of the stock being unsustainably exploited. However, ICES notes that the TAC is not currently controlling exploitation. http://www.ices.dk/sites/pub/Publication\ Reports/Forms/DispForm.aspx?ID=34726.

## Type of assessment

SPALY update of ASAP assessment.

## ICES advice applicable to 2022 and 2023

ICES advises that when the MSY approach and precautionary considerations are applied, there should be zero catches in 2022 and 2023.
https://doi.org/10.17895/ices.advice. 5224

### 37.1 General

## Stock description and management units

The stock and the management unit are both ICES Division 7.a (Irish Sea). Whiting landings taken or reported in ICES rectangles 33E2 and 33E3 have been reassigned to the 7.b,c,e-k whiting stock since 2003.


Management applicable to 2022 and 2023
The minimum conservation reference size of whiting is 27 cm . This stock is subject to the landings obligation as part of the Commission Delegated Regulation (EU) 2018/2034.

In 2022, the TAC was set at 721 t .
In 2021, there was no agreed TAC until mid-year when it was set at 721 t .
In 2020, the TAC was set to 721 t . This followed an ICES technical service that examined the likely catches in 2020 for specific bycatch stocks that have zero catch advice.
https://www.ices.dk/sites/pub/Publication\ Reports/Advice/2019/Special Requests/eu.2019.23.pdf

In 2019 the TAC was set to 727 t . This TAC was later increased to 1246 t following from ICES advice in March 2019.
https://www.ices.dk/sites/pub/Publication\ Reports/Advice/2019/Special Requests/eu.2019.02.pdf

Official landings as reported to ICES in 2021 were 149 t , an increase from 102 t in 2020.

## TAC 2021

| Species: | Whiting <br> Merlangius merlangus | Zone:7 a  <br>  $(\mathrm{WHG} / 07 \mathrm{~A})$. |
| :---: | :---: | :---: |
| Belgium | $2^{(1)}$ | Analytical TAC |
| France | $22^{(1)}$ | Article 9 of this Regulation applies |
| Ireland | $280{ }^{\text {(1) }}$ | Article 3 of Regulation (EC) No 847/96 shall not apply |
| The Netherlands | $0^{(1)}$ | Article 4 of Regulation (EC) No 847/96 shall not apply |
| Union | $305^{(1)}$ |  |
| United Kingdom | $416^{(1)}$ |  |
| TAC | $721{ }^{(1)}$ |  |
| (1) | Exclusively for by-catches of whiting in fisheries for oth | No directed fisheries for whiting are permitted under this quota. |

TAC 2022

| Species: | Whiting <br> Merlangius merlangus |  |  | Zone: | $\begin{aligned} & \text { 7a } \\ & (\mathrm{WHG} / 07 \mathrm{~A} .) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  | 2 | (1) | Analy |  |
| France |  | 22 | (1) | Artic | Regulatio |
| Ireland |  | 274 | ${ }^{(1)}$ | Articl | gulation (EC |
| Netherlands |  | 1 | (1) | appl |  |
| Union |  | 299 | (1) | Articl apply | egulation (EC |
| United Kingdom |  | 422 | (1) |  |  |
| TAC |  | 721 (1) |  |  |  |
| (1) Exclusively for by-catches of whiting in fisheries for other species. No directed fisheries for whiting are permitted under this quota. |  |  |  |  |  |

## Fishery in 2021

The characteristics of the fishery are described in the stock annex.
The fishery in 2021 was prosecuted by the same fleets and gears as in recent years.

The majority of catches are discards are from Nephrops directed fleets. The main fleets landing whiting are fin-fish directed fleets from Ireland and Northern Ireland. In recent years landings were submitted for the PTM_SPF métier. These are likely from trips targeting herring where whiting was a bycatch.

Figure 37.2 shows the contribution of catch by fleet.
Table 37.1 gives the official nominal landings of 7 .a whiting as reported by each country to ICES. Working Group estimates of the landings and discards for the main fleets are given in Table 37.2. In recent years the values provided to the WG are very similar to officially reported landings. Ireland, Belgium and UK(NI) submitted discard estimates for 2021. Total discard estimates were to be 1571 t in 2021.

No BMS landings or logbook registered discards were submitted to ICES for 2021.
The closure of the western Irish Sea to whitefish fishing from mid-February till the end of April, designed to protect cod, was continued in 2021 but is unlikely to have affected whiting catches which are mainly bycatch in the derogated Nephrops fishery. Nephrops vessels can obtain a derogation to fish in certain sections of the closed area, providing they fit separator panels to their nets to allow escape of cod and other fish. The TR2 fleet in 7.a are obliged to use one of four types
of cod selective measures, namely a 'Swedish' grid; the inclined separator panel, SELTRA trawl or 300 square-mesh panel.

A summary of the 2021 catches by main gear types is presented below.

| Catch (2021) | Landings | Discards |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1662 tonnes | Finfish-directed ot- <br> ter trawls | Nephrops-directed otter <br> trawls | Other gears | Nephrops-directed otter <br> trawls | Other gears |
|  | $96.5 \%$ | $0.9 \%$ | $2.6 \%$ | $97.6 \%$ | $2.4 \%$ |
|  |  | 91 tonnes | 1571 tonnes |  |  |

### 37.2 Information from the Industry

There was no information on the whiting stock from the industry.

### 37.3 Data

Data were provided by all countries according to the data call.
For WGCSE (2022) all data have been updated where possible. To allow an age-based assessment, catch numbers-at-age, catch weights-at-age, stock weights-at-age have all be constructed since 2003 (WGCSE, 2017). These updates are documented in the Stock Annex.

## Fishery landings

Working Group estimates of catch available since 1980 are illustrated in Figure 37.1 and indicate the declining trend since the start of the time-series. In 2021, there was a slight increase in landings from 88 t to 91 t .

The introduction of UK and Irish legislation requiring registration of fish buyers and sellers may mean that the reported landings from 2006 onwards are more representative of actual landings.

Working group estimates of landings are corrected for misreporting in the past. There is information that officially reported landings of whiting, especially around the mid-1990s, have been inaccurate due to misreporting. Landings data have previously been partially corrected for by using sample-based estimates of landings at a number of Irish Sea ports. Due to the low level of landings recently, this has not been carried out since 2003. As for 7.a cod and haddock, the whiting landings taken or reported in ICES rectangles 33 E 2 and 33 E 3 have been reassigned to the 7.e-k whiting stock since 2003 (b Based on UK (N.Ireland) and Ireland data.
c Based on data from Ireland.
d Based on data from Northern Ireland.

* Preliminary (and rounded).
e Raised using Days.
f Average IR-OTB discards (2017-2019).
Table 37.3).


## Fishery discards

Discard estimates are available from Northern Ireland and Belgian fleets. Raising methods used are described in the stock annex for 7 .a whiting.

## Landings-at-age data

Landings numbers-at-age are given in Table 37.4. For the 2003 data onwards, the catch and mean weight-at-age are estimated using combined UK (NI) and Irish quarterly length-weight relationships and age-length keys. These data are raised to the international catch data provided to ICES. Typically, quarterly landings are provided by the UK (Scotland), Belgium and France and annual landings are provided by UK (IOM). The quality of the landings-at-age data has been declining in recent years due to reduced sample numbers commensurate with the decline in landings. In 2021, landings at-age were provided by Ireland and Northern Ireland.

Sampling and raising methods previously used are described in the stock annex for 7.a whiting. Methods for estimating quantities and composition of landings are described in the stock annex.

## Discards numbers-at-age data

In 2021, discard sampling numbers at age were available from Northern Ireland and Ireland.
Discard number at age are given in Error! Reference source not found.. Discarding of whiting is high within the Irish Sea. Discard Numbers at age were combined for ages 0 to $6+$ and then raised to the international discards. There has been a high number of age 1 and 2 discarded at the start of the time series with almost all age 1 and 2 discarded later in time series (


The length frequency of discards of national sampled fleets in 2019 is given in Figure 37. This information has not been updated for 2021. More detailed information is available in the stock annex.

## Biological data

The derivation of these parameters and variables is described in the stock annex. The Lorenzen method was used to estimate M. This was derived during WKIrish 2 and investigated during WKIrish 3. Maturity-at-age is knife edge at age 2 . Stock weights were also revised at the benchmark meeting. Stock weights-at-age were derived from the catch weights and then smoothed using a three year moving average. Figure 37. shows the stock weights used. There are strong trends in mean weights-at-age over the time-series with a minimum around 2000s for most ages.

There was a small increase in the mid-2000s but overall mean weights are significantly lower than at the start of the series.

## Survey data used in assessment

Table 37.5 describes the survey data made available to the Working Group.
Survey series for whiting provided to the Working Group are further described in the stock annex for 7.a whiting (Section B.3). Five survey series were available. The inclusion of the different available surveys was tested in a series of preliminary model runs at WKIrish 3.

The three surveys used in the assessment are NIGFS-WIBTS-Q1 (G7144), NIGFS-WIBTS-Q4 (G7655), and NI MIK (I9826).

The 2019 UK (E\&W)-BTS-Q3: Corystes Irish Sea Beam-Trawl Survey data were revised for WGCSE 2022. The 2019 data were updated from 340 to 307 for age 0 and 207 to 186 for age 1.

Figure 37. shows the log standardized indices by cohort of the tuning fleets used in the assessment. There are very little cohort signals in any of the indices. The survey data show a major change in the age structure of the stock around the mid-2000s. The two NI surveys show that older fish disappear around 2003 in the Q1 survey and around 2004 in the Q4 survey. This is mainly due to a decline whiting catches in the Eastern Irish Sea stratum which was explored in detail at WKIrish.

### 37.4 Historical Stock Development

Model used: ASAP
Software used: ASAP V3.0.17 NOAA Fisheries toolbox (http://nft.nefsc.noaa.gov)
FLR with R version 3.6.1 (64-bit) with packages FLEDA 2.5.2, FLCore 2.6.15, FLAssess 2.6.3, and Flash (http://flr-project.org)

## Data screening

The general approach to data screening and analysis was followed in addition to the data exploration tools available in the FLR package FLEDA. The results of the data screening are fully documented using R markdown and are available in the folder 'Data $\backslash$ Whg 7.a $\backslash$ Assessment. on SharePoint. Table 37.6 shows the ASAP input data.

## Final update assessment

The final assessment was run using the same settings as described in WKIrish 3. These final settings are described in the Stock Annex. The exception to this is the CV of 0.3 used for catch numbers-at-age for 2020.

Figure 37. shows the selectivity-at-age in the catch. Full selectivity is assumed for age 3 and the model is allowed to estimate ages 1 and 2 . Table 37.7 shows the model estimates.

The observed and predicted index CPUE values are shown in Figure 37.. There is poor fit to the Northern Irish groundfish survey indices in the first half of the series but it improves in recent years.

The observed and predicted catches are shown in Figure 37.. Fit to the overall catch is reasonably good. There is some deviation in the early to mid-1990s. This is most likely due to the introduction of the survey data into the assessment model.

Figure 37. shows the retrospective analysis. The predicted catch shows no obvious retrospective pattern. The recruitment shows a slight underestimate in the last year. There is some deviation in the early part of the time-series when the surveys were first introduced. However, recent estimates of SSB and F are consistent with no apparent bias.

A Mohn's rho analysis was conducted based on the ASAP stock assessment results, i.e. the last data year (2020) was used as the final year for comparison of SSB, F and recruitment and based on a five-year retrospective analysis. The results from the Mohn's rho analysis are shown in the following table:

|  | SSB | F (ages 1-3) | recruitment |
| :--- | :--- | :--- | :--- |
| Mohn's rho value | 0.187 | -0.121 | 0.27 |

The Mohn's rho values for this assessment are below the threshold imposed by ICES of $20 \%$ for recruitment and $15 \%$ for fishing mortality.

## The state of the stock

Table 37.8 shows the estimated fishing mortality-at-age and Table 37.9 shows the stock numbers-at-age. The stock summary is given in Table 37.10 and Figure 37.

The present stock size is extremely low. SSB has declined since the start of the time-series and has been well below Blim since the mid-1990s. Recruitment has been low since the early 1990s with a slight increase in recent years. Large variations in fishing mortality estimates have been observed in recent years. F has been well above Flim since the early 1990s.

### 37.5 Short-term predictions

Short-term projections were performed using FLR libraries. Recruitment for 2022-2024 was estimated at 119242 (GM 2000-2020:thousands). As the retrospective pattern shows an underestimate of recruitment, the terminal year was excluded from the GM for the WGCSE, 2022 assessment. Three year averages (2019--2021) were used for F (unscaled) and weights-at-age.

Input data for the short-term forecast are given in Table 37.11. The single-option output is given in Table 37.12 , Table 37.13 and Table 37.14 gives the management options.


Figure 37.. The 2019-2020 year class estimates from ASAP accounts for $72 \%$ of the projected landings in 2023. The 2022 GM assumption contributes considerably to the estimated SSB in 2024 as does the 2021 ASAP assessment.

### 37.6 Medium-term projection

There is no analytical assessment for this stock.

### 37.7 MSY evaluations and Biological Reference Points

ICES carried out and evaluation of MSY and PA reference points for this stock at WKIrish 3. The results are summarized below:

|  | Type | Value | Technical basis |
| :---: | :---: | :---: | :---: |
| MSY | MSY $\mathrm{B}_{\text {trigger }}$ | 16300 t | $\mathrm{B}_{\mathrm{pa}}$ |
| Approach | $\mathrm{F}_{\text {MSY }}$ | 0.22 | Median point estimates of EqSim with combined SR |
|  | $\mathrm{F}_{\text {MSY lower }}$ | 0.158 | Median point estimates of EqSim with combined SR |
|  | $\mathrm{F}_{\text {MSY upper }}$ | 0.294 | Median point estimates of EqSim with combined SR |
|  | Blim | 10000 t | Below 10000 t recruitment is impaired |
| Precautionary | $\mathrm{B}_{\text {pa }}$ | 16300 t | $\mathrm{B}_{\text {lim }}$ combined with the assessment error |
| Approach | $\mathrm{F}_{\text {lim }}$ | 0.37 | F with 50\% probability of SSB less than $\mathrm{Bl}_{\mathrm{lim}}$ |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.22 | $\mathrm{F}_{\mathrm{p} .05}$; the F that leads to $S S B \geq \mathrm{Bl}_{\text {lim }}$ with $95 \%$ probability |

In 2021, ICES updated the basis for $\mathrm{F}_{\mathrm{pa}}$ as "the F that leads to $\mathrm{SSB} \geq \mathrm{Blim}$ with $95 \%$ probability", ICES (2021). Prior to this, it was based on "Flim combined with the assessment error", ICES (2017). The $\mathrm{F}_{\mathrm{pa}}$ value of 0.22 remains unchanged.

### 37.8 Management plans

No management plan has been agreed or proposed.

### 37.9 Uncertainties and bias in assessment and forecast

This stock was benchmarked in January 2017. The result of the benchmark was that the stock was elevated from a category 3 stock (trend-based assessment) to a category 1 stock (analytical assessment). The assessment includes information from the commercial fishery, including both landings and discards, and takes into account selectivity changes that have occurred in 1995. Three survey series are used within the assessment. Natural mortality parameters were updated to reflect current stock dynamics. The highly fluctuating estimates of fishing mortality in recent years (2002-present) are likely to be the result of variability in the sampling data and discard estimates. Despite this inherent uncertainty, it is clear from the assessment and additional information from surveys that the stock remains extremely low.

Stock status classification relative to MSY proxies is given below. This has not been updated for WGCSE, 2022.

|  | Fishing pressure |  |  |  |  | Stock size |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 2018 | 2019 |  | 2020 |  | 2019 | 2020 |  | 2021 |
| Maximum sustainable yield | $\mathrm{F}_{\text {MSY }}$ | * | ( | ( | Above | $\begin{aligned} & \text { MSY } \\ & \text { B trigger }^{\text {min }} \end{aligned}$ | $X$ |  |  | Below trigger |
| Precautionary approach | $F_{p a}, F_{\text {lim }}$ | $x$ |  | $x$ | Harvested unsustainably | $\mathrm{B}_{\mathrm{pa}}, \mathrm{B}_{\mathrm{lim}}$ | ( | ( |  | Reduced reproductive capacity |
| Management plan | $\mathrm{F}_{\text {MGT }}$ | - | - | - | Not applicable | $\mathrm{B}_{\mathrm{MGT}}$ | - | - |  | Not applicable |

### 37.10 Recommendations for next benchmark assessment

This stock was benchmarked in 2017 as part of the WKIrish process. A number of recommendations for future work were made and these are listed below. Given the current stock status there is no urgency to schedule another benchmark for this stock in the short term.

## Assessment method

Currently a single fleet ASAP with fixed selection assumption is used. Exploring alternative modelling frameworks which allow for changes in selection should be investigated. There is very little data to inform the question whether survey catchability is flat-topped or dome-shaped. At the moment the highly truncated age structure means that this makes little difference in the model outputs. However, if the stock recovers and more older fish appear then this will need to be revisited.

## Biological parameters

New natural mortality estimates from the Irish Sea EWE model should be included in the assessment. The stock shows very strong changes in weights-at-age over time (they can change by a factor of up to 2). This is likely to affect the natural mortality. Further information to support this would be very useful for future benchmarks.

## Discards

Discards data remain highly uncertain for this stock. This probably contributes to the variable F patterns observed. Partitioning catch data into landings and discards or by fleet with different CVs may help smooth out some of this variability.

## Life-history parameters

Mean weights show trends which are currently smoothed. This should be explored further with a view to improving the approach and possibly using it in forecasts.

## Other issues

Stock identity is assumed to be appropriate but there are east-west differences in population structure and in the past there has been speculation about emigration to 7 g .

## Sampling

Discard sampling should be improved for this stock since discards account for the vast majority of the catch in number. Despite various management initiatives discarding remains sporadic and high in the Nephrops fishery.

## Tuning series

Currently calculated survey CVs are not used in ASAP. It might be worth exploring the impacts of using actual values instead of an assumed fixed CV in future assessment models.

The FSP survey potentially has useful information on the older fish (even though the survey is discontinued). Including the survey in the final assessment run resulted in many of the retrospective runs to fail to converge. It appears therefore that it causes the model to be unstable and was omitted from the final run. For future benchmarks it may be useful to investigate why this survey makes the model unstable.

### 37.11 Management considerations

Discarding in the Nephrops fishery is the main management issue. Despite the implementation of several technical measures, which experimentally reduce whiting catches, as part of the cod longterm management plan and the full implementation of the landings obligation in 2019, the discards estimates still remain high, ca. 1089 t . This stock is a major 'choke species' for the 7.a Nephrops fishery in the context of the landing obligation.

Effort limitations are in force within the Irish Sea as a result of the cod long-term management plan. These effort limitations have not significant reduced mortality on whiting.

Whiting has a low market value, which is likely to contribute to discarding rates.
Technical measures applied to this stock include a minimum conservation reference size $(\geq 27 \mathrm{~cm})$, whiting now mature well below this MCRS.

### 37.12 References

ICES, 2017 :Report of the Benchmark Workshop on the Irish Sea Ecosystem (WKIrish3), 30 January-3 February 2017, Galway, Ireland, ICES CM 2017/BSG:01.

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### 37.13 Tables

Table 37.1. Official landings ( t ) of Whiting in Division 7.a as reported to ICES.

| $\begin{aligned} & \frac{1}{\pi} \\ & \stackrel{\pi}{0} \end{aligned}$ |  | $\begin{aligned} & \text { 쓴 } \\ & \text { 뀬 } \end{aligned}$ | $\begin{aligned} & \text { ס } \\ & \text { 들 } \\ & \underline{\underline{N}} \end{aligned}$ |  |  | $\begin{aligned} & \text { 듣 } \\ & \text { io } \end{aligned}$ |  |  | $\underset{ }{〕}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 90 | 1,063 | 4,394 |  | 5,823 |  | 15 | 107 |  | 11,492 |
| 1989 | 92 | 533 | 3,871 |  | 6,652 |  | 26 | 154 |  | 11,328 |
| 1990 | 142 | 528 | 2,000 |  | 5,202 |  | 75 | 236 |  | 8,183 |
| 1991 | 53 | 611 | 2,200 |  | 4,250 |  | 74 | 223 |  | 7,411 |


| 1992 | 78 | 509 | 2,100 |  | 4,089 |  | 44 | 274 |  | 7,094 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 50 | 255 | 1,440 |  | 3,859 |  | 55 | 318 |  | 5,977 |
| 1994 | 80 | 163 | 1,418 |  | 3,724 |  | 44 | 208 |  | 5,637 |
| 1995 | 92 | 169 | 1,840 |  | 3,125 |  | 41 | 198 |  | 5,465 |
| 1996 | 80 | 78 | 1,773 | 17 | 3,557 |  | 28 | 48 |  | 5,581 |
| 1997 | 47 | 86 | 1,119 | 14 | 3,152 |  | 24 | 30 |  | 4,472 |
| 1998 | 52 | 81 | 1,260 | 7 | 1,900 |  | 33 | 22 |  | 3,355 |
| 1999 | 46 | 150 | 509 | 6 | 1,229 |  | 5 | 44 |  | 1,989 |
| 2000 | 30 | 59 | 353 | 1 | 670 |  | 2 | 15 |  | 1,130 |
| 2001 | 27 | 25 | 482 |  | 506 |  | 1 | 25 |  | 1,066 |
| 2002 | 22 | 33 | 347 |  | 284 |  | 1 | 27 |  | 714 |
| 2003 | 13 | 29 | 265 |  | 130 | 85 | 1 | 31 |  | 554 |
| 2004 | 11 | 8 | 96 |  | 82 |  | 1 | 6 |  | 204 |
| 2005 | 10 | 13 | 94 |  | 47 |  |  | <0.5 |  | 164 |
| 2006 | 4 | 4 | 55 |  | 22 |  |  | <0.5 |  | 85 |
| 2007 | 3 | 3 | 187 |  | 3 |  | 1 | <0.5 |  | 197 |
| 2008 | 2 | 2 | 68 |  | 11 |  | 1 |  |  | 84 |
| 2009 | 2 |  | 78 |  | 20 |  |  |  |  | 100 |
| 2010 | 5 | 3 | 97 |  | 16 |  | <0.5 |  |  | 121 |
| 2011 | 4 | 3 | 95 |  | 16 |  | <0.5 |  |  | 118 |
| 2012 | 5 | 1 | 58 |  | 10 |  |  | 1 | 11 | 86 |
| 2013 | 2 | $<0.5$ | 44 |  |  |  | <0.1 | 2 | 20 | 68 |
| 2014 | 2 | <0.5 | 60 |  | 11 |  | <0.1 |  |  | 73 |
| 2015 | 1 | $<0.5$ | 49 |  | 8 |  |  |  |  | 59 |
| 2016 | 1 | $<0.5$ | 44 |  | 5 |  | <0.1 |  |  | 50 |
| 2017 | 2 | $<0.5$ | 32 |  | 17 |  | <0.1 |  |  | 50 |
| 2018 | 1 |  | 44 |  | 19 |  | <0.5 |  |  | 63 |
| 2019 | 4 |  | 129 |  | 63 |  | <0.1 |  |  | 196 |
| 2020 | 5 | $<0.1$ | 56 |  | 42 |  | <0.1 |  |  | 102 |
| 2021 | 2 | <0.1 | 109 |  | 38 |  |  |  |  | 149 |

* Preliminary.

Table 37.2. ICES estimates of discards, landings and catch of whiting in Division 7.a.

| Year | Discards by Country/Fleet |  |  |  |  | Discards | Landings | Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Nephrops fishery ${ }^{\text {b }}$ | IR-OTB <br> fleet ${ }^{\text {ce }}$ | NI Nephrops fishery ${ }^{\text {d }}$ | Belgium | UK (E\&W) <br> fleet |  |  |  |
| 1988 | 1,611 |  |  |  |  | 1,611 | 10,245 | 11,856 |
| 1989 | 2,103 |  |  |  |  | 2,103 | 11,305 | 13,408 |
| 1990 | 2,444 |  |  |  |  | 2,444 | 8,212 | 10,656 |
| 1991 | 2,598 |  |  |  |  | 2,598 | 7,348 | 9,946 |
| 1992 | 4,203 |  |  |  |  | 4,203 | 8,588 | 12,791 |
| 1993 | 2,707 |  |  |  |  | 2,707 | 6,523 | 9,230 |
| 1994 | 1,173 |  |  |  |  | 1,173 | 6,763 | 7,936 |
| 1995 | 2,151 |  |  |  |  | 2,151 | 4,893 | 7,044 |


| 1996 | 3,631 |  |  |  |  | 3,631 | 4,335 | 7,966 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 1,928 |  |  |  |  | 1,928 | 2,277 | 4,205 |
| 1998 | 1,304 |  |  |  |  | 1,304 | 2,229 | 3,533 |
| 1999 | 1,092 |  |  |  |  | 1,092 | 1,670 | 2,762 |
| 2000 | 2,118 |  |  |  |  | 2,118 | 762 | 2,880 |
| 2001 | 1,012 |  |  |  |  | 1,012 | 733 | 1,745 |
| 2002 | 740 |  |  |  |  | 740 | 747 | 1,487 |
| 2003 |  | 480 |  |  |  | 480 | 517 | 996 |
| 2004 |  | 905 |  |  |  | 905 | 133 | 1,038 |
| 2005 |  | 272 |  |  |  | 272 | 125 | 397 |
| 2006 |  | 1,580 | 193 |  |  | 1,773 | 64 | 1,837 |
| 2007 |  | 725 | 787 |  |  | 1,512 | 35 | 1,547 |
| 2008 |  | 693 | 476 |  |  | 1,169 | 37 | 1,206 |
| 2009 |  | 688 | 633 |  |  | 1,321 | 39 | 1,360 |
| 2010 |  | 240 | 914 |  |  | 1,154 | 30 | 1,184 |
| 2011 |  | 330 | 616 |  |  | 946 | 31 | 977 |
| 2012 |  | 257 | 1,065 | 17 | 1 | 1,339 | 60 | 1,399 |
| 2013 |  | 95 | 833 | 17 | 3 | 948 | 33 | 981 |
| 2014 |  | 263 | 1,645 | 15 | 28 | 1,951 | 23 | 1,974 |
| 2015 |  | 438 | 1,074 | 9 | 1 | 1,521 | 28 | 1,549 |
| 2016 |  | 173 | 589 |  | 3 | 765 | 15 | 780 |
| 2017 |  | 122 | 544 |  | 1 | 667 | 36 | 703 |
| 2018 |  | 98 | 754 |  | $<0.5$ | 853 | 46 | 899 |
| 2019 |  | 86 | 897 | 20 | 87 | 1,089 | 172 | 1,261 |
| 2020 |  | $10)^{\text {f }}$ | 906 | 22 | Na | 1,030 | 88 | 1,118 |
| 2021 |  | 431 | 1,118 | 22 |  | 1,571 | 81 | 1,662 |

b Based on UK (N.Ireland) and Ireland data.
c Based on data from Ireland.
d Based on data from Northern Ireland.

* Preliminary (and rounded).
e Raised using Days.
f Average IR-OTB discards (2017-2019).

Table 37.3. Whiting landings taken or reported in ICES rectangles $33 E 2,33 E 3$ and $33 E 4$ have been reassigned to the 7.ek whiting stock since 2003.

| Year | Official landings | ICES landings | ICES Discards | ICES catch | Landings taken or reported in <br> rectangles 33E2 and 33E3 |
| :--- | :---: | :---: | :---: | :---: | :---: |
| 1988 | 11,492 | 10,245 | 1,611 | 11,856 |  |
| 1989 | 11,328 | 11,305 | 2,103 | 13,408 |  |
| 1990 | 8,183 | 8,212 | 2,444 | 10,656 |  |
| 1991 | 7,411 | 7,348 | 2,598 | 9,946 |  |
| 1992 | 5,977 | 8,588 | 4,203 | 12,791 |  |
| 1993 | 6,523 | 9,707 |  |  |  |


| Year | Official landings | ICES landings | ICES Discards | ICES catch | Landings taken or reported in rectangles $33 E 2$ and 33E3 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 5,637 | 6,763 | 1,173 | 7,936 |  |
| 1995 | 5,465 | 4,893 | 2,151 | 7,044 |  |
| 1996 | 5,581 | 4,335 | 3,631 | 7,966 |  |
| 1997 | 4,472 | 2,277 | 1,928 | 4,205 |  |
| 1998 | 3,355 | 2,229 | 1,304 | 3,533 |  |
| 1999 | 1,989 | 1,670 | 1,092 | 2,762 |  |
| 2000 | 1,130 | 762 | 2,118 | 2,880 |  |
| 2001 | 1,066 | 733 | 1,012 | 1,745 |  |
| 2002 | 714 | 747 | 740 | 1,487 |  |
| 2003 | 554 | 517 | 480 | 996 | 159 |
| 2004 | 204 | 133 | 905 | 1,038 | 51 |
| 2005 | 164 | 125 | 272 | 397 | 33 |
| 2006 | 85 | 64 | 1,773 | 1,837 | 22 |
| 2007 | 197 | 35 | 1,512 | 1,547 | 161 |
| 2008 | 84 | 37 | 1,169 | 1,206 | 44 |
| 2009 | 100 | 39 | 1,321 | 1,360 | 63 |
| 2010 | 121 | 30 | 1,154 | 1,184 | 91 |
| 2011 | 118 | 31 | 946 | 977 | 75 |
| 2012 | 86 | 60 | 1,339 | 1,399 | 43 |
| 2013 | 68 | 33 | 948 | 981 | 33 |
| 2014 | 73 | 23 | 1,951 | 1,974 | 50 |
| 2015 | 59 | 28 | 1,521 | 1,549 | 34 |
| 2016 | 50 | 15 | 765 | 780 | 40 |
| 2017 | 50 | 36 | 667 | 703 | 20 |
| 2018 | 63 | 46 | 853 | 899 | 18 |
| 2019 | 196 | 172 | 1,089 | 1,261 | 24 |
| 2020 | 102 | 88 | 1,030* | 1,118 | 14 |
| 2021 | 149 | 91 | 1,571 | 1,662 | 59 |

[^22]Table 37.4. Whiting7.a. Landings numbers-at-age.

|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0 | 14520 | 21811 | 6468 | 2548 | 350 | 0 |
| 1981 | 0 | 11203 | 29011 | 16004 | 2596 | 821 | 0 |
| 1982 | 41 | 5427 | 18098 | 19340 | 6108 | 813 | 0 |
| 1983 | 0 | 4886 | 9943 | 9100 | 4530 | 1165 | 321 |
| 1984 | 0 | 18254 | 12683 | 5257 | 2571 | 1045 | 402 |
| 1985 | 0 | 15540 | 35324 | 8687 | 996 | 0 | 675 |
| 1986 | 0 | 6306 | 16839 | 10809 | 1877 | 285 | 0 |
| 1987 | 0 | 10149 | 21563 | 6968 | 1943 | 242 | 0 |
| 1988 | 0 | 6983 | 25768 | 6989 | 1513 | 396 | 0 |
| 1989 | 0 | 11645 | 14029 | 13011 | 3645 | 490 | 0 |
| 1990 | 0 | 9502 | 17604 | 4734 | 1477 | 318 | 0 |
| 1991 | 102 | 7426 | 18406 | 5829 | 993 | 0 | 311 |
| 1992 | 0 | 8380 | 21907 | 7959 | 1374 | 462 | 0 |
| 1993 | 38 | 2742 | 21468 | 7327 | 932 | 0 | 135 |
| 1994 | 0 | 3245 | 6983 | 18509 | 1801 | 208 | 0 |
| 1995 | 0 | 1124 | 10095 | 3020 | 4444 | 233 | 0 |
| 1996 | 129 | 1652 | 6162 | 7432 | 1263 | 1082 | 135 |
| 1997 | 0 | 610 | 4239 | 2567 | 1795 | 87 | 79 |
| 1998 | 0 | 329 | 3287 | 4727 | 888 | 261 | 95 |
| 1999 | 1 | 341 | 2806 | 2607 | 741 | 160 | 119 |
| 2000 | 0 | 319 | 1364 | 1002 | 299 | 115 | 15 |
| 2001 | 0 | 111 | 1189 | 1006 | 171 | 53 | 20 |
| 2002 | 0 | 67 | 748 | 1480 | 376 | 48 | 41 |
| 2003 | 0 | 89 | 1051 | 606 | 199 | 0 | 0 |
| 2004 | 0 | 0 | 17 | 117 | 150 | 17 | 0 |
| 2005 | 0 | 0 | 101 | 216 | 95 | 21 | 3 |
| 2006 | 0 | 34 | 41 | 88 | 39 | 9 | 1 |
| 2007 | 0 | 24 | 41 | 32 | 10 | 3 | 0 |
| 2008 | 0 | 38 | 66 | 25 | 5 | 1 | 0 |


|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 0 | 65 | 44 | 22 | 4 | 1 | 0 |
| 2010 | 0 | 18 | 83 | 11 | 3 | 0 | 0 |
| 2011 | 0 | 1 | 17 | 59 | 15 | 3 | 0 |
| 2012 | 0 | 4 | 29 | 80 | 60 | 9 | 1 |
| 2013 | 8 | 81 | 36 | 20 | 5 | 1 | 1 |
| 2014 | 0 | 2 | 25 | 24 | 11 | 1 | 1 |
| 2015 | 0 | 2 | 25 | 24 | 11 | 1 | 1 |
| 2016 | 0 | 0 | 6 | 21 | 10 | 3 | 0 |
| 2017 | 0 | 0 | 9 | 50 | 43 | 5 | 1 |
| 2018 | 0 | 1 | 14 | 70 | 38 | 19 | 2 |
| 2019 | 0 | 0 | 146 | 181 | 72 | 45 | 23 |
| 2020 | 0 | 0 | 58 | 138 | 93 | 18 | 10 |
| 2021 | 0 | 0 | 32 | 119 | 62 | 42 | 6 |

Table 37.5. Whiting7.a. Discards numbers-at-age.

|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 12786 | 32318 | 6888 | 65 | 26 | 0 | 0 |
| 1981 | 9865 | 24935 | 9162 | 162 | 26 | 0 | 0 |
| 1982 | 4047 | 8489 | 560 | 19 | 0 | 0 | 0 |
| 1983 | 23847 | 7328 | 2036 | 9 | 0 | 0 | 0 |
| 1984 | 26394 | 33900 | 1568 | 11 | 0 | 0 | 0 |
| 1985 | 12380 | 26461 | 1859 | 9 | 0 | 0 | 0 |
| 1986 | 28364 | 21111 | 1464 | 33 | 0 | 0 | 0 |
| 1987 | 16594 | 40598 | 1875 | 0 | 0 | 0 | 0 |
| 1988 | 6922 | 17958 | 1940 | 0 | 0 | 0 | 0 |
| 1989 | 17247 | 20701 | 2476 | 26 | 0 | 0 | 0 |
| 1990 | 4216 | 31810 | 3353 | 72 | 0 | 0 | 0 |
| 1991 | 20349 | 29334 | 3823 | 146 | 1 | 0 | 0 |
| 1992 | 1497 | 61451 | 10404 | 97 | 0 | 0 | 0 |
| 1993 | 12639 | 13979 | 17707 | 426 | 5 | 0 | 0 |
| 1994 | 3731 | 12063 | 1812 | 1702 | 29 | 0 | 0 |
| 1995 | 7118 | 17613 | 7015 | 492 | 234 | 0 | 0 |
| 1996 | 12732 | 39647 | 8168 | 1976 | 81 | 0 | 0 |
| 1997 | 8163 | 25497 | 5352 | 689 | 141 | 0 | 0 |
| 1998 | 6096 | 27131 | 2293 | 550 | 44 | 0 | 0 |
| 1999 | 20851 | 7677 | 2117 | 228 | 34 | 2 | 2 |
| 2000 | 7321 | 38922 | 4395 | 564 | 55 | 1 | 10 |
| 2001 | 16940 | 12631 | 3150 | 102 | 10 | 0 | 0 |
| 2002 | 8538 | 13412 | 1588 | 231 | 33 | 0 | 1 |
| 2003 | 12389 | 4595 | 201 | 0 | 0 | 0 | 0 |
| 2004 | 19699 | 14938 | 345 | 59 | 0 | 0 | 0 |
| 2005 | 643 | 5797 | 346 | 16 | 3 | 0 | 0 |
| 2006 | 15764 | 20590 | 613 | 21 | 0 | 0 | 0 |
| 2007 | 17436 | 24319 | 747 | 50 | 0 | 0 | 0 |
| 2008 | 10645 | 19994 | 676 | 16 | 0 | 0 | 0 |


|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 6622 | 27448 | 1176 | 0 | 0 | 0 | 0 |
| 2010 | 3946 | 15102 | 2810 | 64 | 1 | 0 | 0 |
| 2011 | 25982 | 8197 | 658 | 314 | 0 | 0 | 0 |
| 2012 | 6637 | 31020 | 790 | 37 | 1 | 3 | 0 |
| 2013 | 8493 | 11945 | 613 | 4 | 0 | 0 | 0 |
| 2014 | 13467 | 27553 | 2425 | 259 | 10 | 0 | 0 |
| 2015 | 3883 | 23595 | 2603 | 223 | 1 | 0 | 0 |
| 2016 | 4509 | 5780 | 4804 | 294 | 15 | 0 | 0 |
| 2017 | 3559 | 5870 | 4385 | 240 | 14 | 0 | 0 |
| 2018 | 6523 | 7386 | 2557 | 614 | 92 | 10 | 0 |
| 2019 | 6429 | 14041 | 3986 | 571 | 57 | 7 | 0 |
| 2020 | 11987 | 26870 | 978 | 50 | 3 | 0 | 0 |
| 2021 | 4272 | 18880 | 6496 | 396 | 18 | 4 | 0 |

Table 37.5. Whiting in 7.a. Survey data available.

NIGFS-WIBTS-Q1: Northern Ireland March Groundfish Survey

| $\begin{gathered} 1993 \\ \hline 1 \end{gathered}$ | 2021 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0.21 | 0.25 |  |  |  |
| 1 | 6 |  |  |  |  |  |
| 1 | 665.6 | 710.3 | 81.2 | 11.7 | 4.3 | 0.8 |
| 1 | 1804.6 | 262.1 | 299.2 | 44.7 | 11.9 | 8.1 |
| 1 | 1688.9 | 635.7 | 174.2 | 88.4 | 22.0 | 6.3 |
| 1 | 1468.4 | 334.0 | 213.0 | 35.1 | 37.2 | 5.4 |
| 1 | 1406.1 | 1536.4 | 156.0 | 52.8 | 4.5 | 13.7 |
| 1 | 1485.0 | 754.4 | 415.4 | 29.7 | 7.4 | 1.8 |
| 1 | 1369.4 | 373.2 | 111.2 | 41.5 | 3.7 | 1.0 |
| 1 | 2302.4 | 410.9 | 181.8 | 26.6 | 3.7 | 0.0 |
| 1 | 1065.7 | 696.5 | 124.6 | 13.7 | 5.9 | 2.7 |
| 1 | 2307.7 | 686.7 | 175.3 | 52.9 | 11.2 | 1.4 |
| 1 | 1495.1 | 905.2 | 130.2 | 10.9 | 1.6 | 0.1 |
| 1 | 1609.8 | 231.7 | 61.4 | 2.7 | 1.3 | 0.2 |
| 1 | 689.3 | 124.0 | 28.5 | 12.3 | 2.8 | 0.1 |
| 1 | 959.8 | 235.6 | 30.3 | 6.0 | 0.1 | 0.1 |
| 1 | 905.0 | 158.6 | 14.9 | 2.7 | 0.2 | 0.0 |
| 1 | 756.7 | 347.0 | 45.0 | 2.8 | 0.3 | 0.4 |
| 1 | 1062.3 | 281.1 | 36.3 | 1.8 | 0.2 | 0.1 |
| 1 | 739.4 | 545.8 | 51.6 | 4.7 | 6.4 | 0.0 |
| 1 | 586.4 | 156.5 | 36.0 | 3.9 | 0.6 | 0.0 |
| 1 | 972.2 | 354.4 | 42.3 | 5.9 | 1.2 | 0.0 |
| 1 | 629.6 | 649.3 | 66.7 | 3.5 | 0.5 | 0.0 |
| 1 | 922.1 | 367.6 | 67.0 | 4.3 | 0.2 | 0.1 |
| 1 | 2797.3 | 469.3 | 18.8 | 2.3 | 0.0 | 0.0 |
| 1 | 1409.1 | 924.8 | 38.7 | 1.5 | 0.1 | 0.1 |
| 1 | 888.1 | 831.8 | 142.2 | 11.2 | 0.7 | 0.1 |
| 1 | 431.4 | 296.8 | 119.4 | 17.9 | 2.3 | 0.0 |
| 1 | 568.0 | 831.9 | 347.2 | 43.2 | 6.2 | 0.5 |
| 1 | 1573.5 | 583.4 | 127.3 | 9.2 | 0.3 | 0.6 |
| 1 | 569.4 | 951.5 | 86.2 | 9.9 | 2.9 | 0.6 |

NIGFS-WIBTS-Q4: Northern Ireland October Groundfish Survey

| $\frac{1993}{1}$ | 2021 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0.83 | 0.88 |  |  |  |  |
| 0 | 6 |  |  |  |  |  |  |
| 1 | 714.0 | 1040.5 | 475.9 | 67.5 | 8.2 | 3.1 | 0.3 |
| 1 | 1113.1 | 1320.0 | 208.6 | 150.7 | 33.9 | 2.3 | 0.5 |
| 1 | 3124.4 | 477.3 | 166.5 | 30.6 | 35.6 | 5.4 | 1.2 |
| 1 | 2306.2 | 591.2 | 134.4 | 52.4 | 10.5 | 7.0 | 1.3 |
| 1 | 2626.5 | 676.6 | 497.6 | 61.0 | 18.2 | 4.6 | 4.5 |
| 1 | 2863.5 | 466.8 | 153.8 | 72.8 | 6.2 | 2.2 | 0.1 |
| 1 | 2478.4 | 1079.7 | 192.0 | 51.7 | 43.3 | 3.7 | 1.8 |
| 1 | 2374.3 | 1084.7 | 126.0 | 20.0 | 16.9 | 6.0 | 2.7 |
| 1 | 6356.4 | 658.3 | 270.8 | 28.9 | 4.9 | 2.3 | 0.0 |
| 1 | 2692.4 | 1322.5 | 268.3 | 41.6 | 4.5 | 1.2 | 0.0 |
| 1 | 4431.0 | 1572.3 | 921.1 | 74.8 | 16.8 | 1.5 | 0.0 |
| 1 | 4457.1 | 699.6 | 268.3 | 113.8 | 4.4 | 1.9 | 0.0 |
| 1 | 2377.2 | 487.8 | 183.3 | 15.8 | 1.5 | 0.4 | 0.0 |
| 1 | 2849.2 | 144.8 | 46.8 | 7.9 | 1.8 | 0.0 | 0.0 |
| 1 | 2163.1 | 957.6 | 149.1 | 16.7 | 4.8 | 4.3 | 0.2 |
| 1 | 4884.6 | 1312.6 | 114.3 | 3.8 | 0.2 | 0.0 | 0.0 |
| 1 | 2246.5 | 510.8 | 71.7 | 7.5 | 1.6 | 0.0 | 0.2 |
| 1 | 2274.4 | 312.1 | 259.6 | 8.2 | 0.7 | 0.2 | 0.0 |
| 1 | 3534.1 | 348.4 | 139.7 | 26.3 | 3.5 | 0.9 | 0.0 |
| 1 | 1330.9 | 402.5 | 134.7 | 19.5 | 6.2 | 0.1 | 0.0 |
| 1 | 7135.8 | 354.7 | 155.9 | 31.1 | 1.5 | 0.5 | 0.9 |
| 1 | 4504.0 | 507.7 | 135.5 | 8.8 | 0.7 | 0.0 | 0.0 |
| 1 | 2802.4 | 891.0 | 115.2 | 6.3 | 0.7 | 0.0 | 0.0 |
| 1 | 2718.7 | 859.3 | 203.5 | 31.7 | 3.5 | 0.4 | 0 |
| 1 | 3011.1 | 714.1 | 368.4 | 78.4 | 4.2 | 0.0 | 0.1 |
| 1 | 4424.7 | 897.5 | 367.6 | 23.4 | 8.3 | 0.2 | 0.04 |
| 1 | 5613.5 | 643.2 | 148.5 | 27.4 | 3.2 | 0.3 | 0.00 |
| 1 | 2416.2 | 1157.8 | 98.4 | 16.0 | 0.2 | 0.5 | 0.00 |
| 1 | 5376.7 | 1018.7 | 143.1 | 25.6 | 4.9 | 0.0 | 0.1 |


| 1994 | 2019 |  |  |
| :---: | :---: | :---: | :---: |
| 1 | 1 | 0.46 | 0.5 |
| 0 | 0 |  |  |
| 1 | 778 | 1994 |  |
| 1 | 225 | 1995 |  |
| 1 | 397 | 1996 |  |
| 1 | 205 | 1997 |  |
| 1 | 59 | 1998 |  |
| 1 | 91 | 1999 |  |
| 1 | 40 | 2000 |  |
| 1 | 167 | 2001 |  |
| 1 | 19 | 2002 |  |
| 1 | 148 | 2003 |  |
| 1 | 101 | 2004 |  |
| 1 | 135 | 2005 |  |
| 1 | 118 | 2006 |  |
| 1 | 82 | 2007 |  |
| 1 | 99 | 2008 |  |
| 1 | 173 | 2009 |  |
| 1 | 78 | 2010 |  |
| 1 | 122.2 | 2011 |  |
| 1 | 123.9 | 2012 |  |
| 1 | 197.6 | 2013 |  |
| 1 | 54.9 | 2014 |  |
| 1 | 59.5 | 2015 |  |
| 1 | 6.7 | 2016 |  |
| 1 | 175.45 | 2017 |  |
| 1 | 90.74 | 2018 |  |
| 1 | 164.42 | 2019 |  |
| 1 | N/A | 2020 |  |
| 1 | 108.4 | 2021 |  |

UK (E\&W)-BTS-Q3: Corystes Irish Sea Beam-Trawl Survey - Prime stations only - Effort and numbers-at-age (per km towed)

| 1988 | 2019 |  |  |
| :---: | :---: | :---: | :---: |
| 1 | 1 | 0.75 | 0.79 |
| 0 | 1 |  |  |
| 1 | 96 | 26 | 1988 |
| 1 | 93 | 21 | 1989 |
| 1 | 99 | 33 | 1990 |
| 1 | 216 | 25 | 1991 |
| 1 | 405 | 206 | 1992 |
| 1 | 253 | 95 | 1993 |
| 1 | 205 | 125 | 1994 |
| 1 | 1949 | 87 | 1995 |
| 1 | 169 | 194 | 1996 |
| 1 | 409 | 254 | 1997 |
| 1 | 893 | 199 | 1998 |
| 1 | 550 | 137 | 1999 |
| 1 | 320 | 122 | 2000 |
| 1 | 585 | 195 | 2001 |
| 1 | 280 | 96 | 2002 |
| 1 | 456 | 229 | 2003 |
| 1 | 917 | 330 | 2004 |
| 1 | 849 | 294 | 2005 |
| 1 | 1010 | 228 | 2006 |
| 1 | 339 | 89 | 2007 |
| 1 | 780 | 72 | 2008 |
| 1 | 389 | 371 | 2009 |
| 1 | 324 | 33 | 2010 |
| 1 | 1002 | 341 | 2011 |
| 1 | 442 | 426 | 2012 |
| 1 | 1535 | 228 | 2013 |


| 1988 | 2019 |  |  |
| :--- | :--- | :--- | :--- |
| 1 | 261 | 113 | 2014 |
| 1 | 211 | 213 | 2015 |
| 1 | 666 | 230 | 2016 |
| 1 | 489 | 380 | 2017 |
| 1 | 662 | 186 | 2018 |
| 1 | 307 | $\mathrm{~N} / \mathrm{A}$ | 2019 |
| 1 | $\mathrm{~N} / \mathrm{A}$ | 132 | 2020 |
| 1 | 340 | 2021 |  |

Eastern Irish Sea FSP: Isadale 2005-2013: Numbers of fish per hour towed

| 2005 | 2013 |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 1 | 0.2 | 0.2 |  |  |  |  |
| 1 | 6.0 |  |  |  |  |  |  |
| 1 | 0.2 | 11.1 | 21.1 | 5.3 | 1.0 | 0.0 | 0.7 |
| 1 | 8.7 | 46.7 | 15.2 | 1.9 | 0.5 | 0.0 | 0.0 |
| 1 | 4.2 | 10.8 | 5.6 | 1.0 | 0.3 | 0.0 | 0.0 |
| 1 | 3.7 | 10.3 | 8.6 | 2.0 | 0.4 | 0.3 | 0.0 |
| 1 | 27.3 | 84.9 | 48.7 | 3.6 | 0.3 | 0.0 | 0.0 |
| 1 | 4.5 | 57.9 | 43.5 | 5.0 | 0.2 | 0.1 | 0.0 |
| 1 | 2.2 | 8.4 | 31.9 | 5.1 | 1.0 | 0.0 | 0.0 |
| 1 | 5.2 | 80.9 | 29.8 | 22.1 | 1.2 | 0.1 | 0.0 |
| 1 | 4.2 | 47.4 | 26.4 | 3.1 | 1.7 | 0.0 | 0.0 |

Table 37.6. Whiting 7.a. ASAP input data.

```
# ASAP VERSION 3.0
# Irish Sea Whiting
#
# ASAP GUI 15 AUG 2012
#
# Number of Years
4 2
# First Year
1980
# Number of Ages
7
# Number of Fleets
1
# Number of Sensitivity Blocks
2
# Number of Available Survey Indices
5
# Natural Mortality
\begin{tabular}{lllllll}
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
1.078 & 0.803 & 0.718 & 0.608 & 0.554 & 0.518 & 0.518 \\
\hline .7 & 0 & & & & \\
\hline
\end{tabular}
# Fecundity Option
O
# Fraction of year that elapses prior to SSB calculation (0=Jan-1)
O
```

| \# Maturity |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| 0 | 0 | 1 | 11 | 1 | 1 |  |  |
| \# Number of Weights-at-Age Matrices |  |  |  |  |  |  |  |
| 2 |  |  |  |  |  |  |  |
| \# Weight Matrix-1 |  |  |  |  |  |  |  |
| 0 |  | 0.11 | 0.235 | 0.363 | 0.529 | 0.63 | 0.772 |
| 0.04 |  | 0.118 | 0.24 | 0.364 | 0.529 | 0.63 | 0.888 |
| 0.031 |  | 0.135 | 0.265 | 0.365 | 0.533 | 0.63 | 0.736 |
| 0.033 |  | 0.146 | 0.256 | 0.397 | 0.491 | 0.605 | 0.655 |
| 0.032 |  | 0.125 | 0.244 | 0.403 | 0.55 | 0.7 | 0.745 |
| 0.021 |  | 0.107 | 0.245 | 0.333 | 0.478 | 0.567 | 0.642 |
| 0.025 |  | 0.1 | 0.217 | 0.342 | 0.512 | 0.709 | 0.94 |
| 0.024 |  | 0.101 | 0.217 | 0.363 | 0.535 | 0.72 | 0.933 |
| 0.021 |  | 0.088 | 0.201 | 0.33 | 0.547 | 0.763 | 1.005 |
| 0.026 |  | 0.111 | 0.193 | 0.269 | 0.433 | 0.68 | 1.079 |
| 0.036 |  | 0.094 | 0.204 | 0.31 | 0.436 | 0.676 | 0.8 |
| 0.031 |  | 0.077 | 0.194 | 0.263 | 0.352 | 0.453 | 0.692 |
| 0.014 |  | 0.063 | 0.17 | 0.272 | 0.361 | 0.513 | 1.007 |
| 0.029 |  | 0.067 | 0.142 | 0.228 | 0.331 | 0.454 | 0.892 |
| 0.03 |  | 0.074 | 0.183 | 0.221 | 0.301 | 0.378 | 0.496 |
| 0.031 |  | 0.063 | 0.179 | 0.257 | 0.326 | 0.551 | 1.32 |
| 0.027 |  | 0.057 | 0.159 | 0.23 | 0.284 | 0.364 | 0.715 |
| 0.026 |  | 0.044 | 0.153 | 0.222 | 0.287 | 0.396 | 0.679 |
| 0.017 |  | 0.035 | 0.156 | 0.228 | 0.268 | 0.35 | 0.421 |
| 0.028 |  | 0.044 | 0.161 | 0.246 | 0.324 | 0.351 | 0.325 |


| 0.024 | 0.038 | 0.127 | 0.218 | 0.291 | 0.347 | 0.31 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.017 | 0.036 | 0.132 | 0.301 | 0.338 | 0.538 | 0.337 |
| 0.016 | 0.033 | 0.124 | 0.253 | 0.339 | 0.449 | 0.425 |
| 0.02 | 0.048 | 0.232 | 0.295 | 0.259 | 0 | 0 |
| 0.017 | 0.034 | 0.131 | 0.324 | 0.509 | 0.466 | 0 |
| 0.017 | 0.037 | 0.148 | 0.263 | 0.363 | 0.36 | 0.32 |
| 0.017 | 0.069 | 0.152 | 0.268 | 0.361 | 0.36 | 0.32 |
| 0.023 | 0.042 | 0.122 | 0.295 | 0.434 | 0.624 | 1.26 |
| 0.022 | 0.044 | 0.118 | 0.262 | 0.374 | 0.834 | 1.354 |
| 0.023 | 0.039 | 0.094 | 0.34 | 0.323 | 0.543 | 0 |
| 0.02 | 0.048 | 0.125 | 0.256 | 0.401 | 0.375 | 0 |
| 0.018 | 0.044 | 0.104 | 0.196 | 0.405 | 0.462 | 0.799 |
| 0.023 | 0.035 | 0.109 | 0.275 | 0.398 | 0.41 | 0.305 |
| 0.03 | 0.052 | 0.112 | 0.24 | 0.346 | 0.28 | 0.38 |
| 0.03 | 0.042 | 0.133 | 0.226 | 0.425 | 0.659 | 1.012 |
| 0.022 | 0.044 | 0.127 | 0.291 | 0.448 | 0.298 | 0.482 |
| 0.022 | 0.035 | 0.085 | 0.195 | 0.341 | 0.466 | 0.882 |
| 0.028 | 0.032 | 0.075 | 0.198 | 0.362 | 0.432 | 0.5 |
| 0.021 | 0.045 | 0.104 | 0.161 | 0.24 | 0.319 | 0.408 |
| 0.02 | 0.033 | 0.104 | 0.175 | 0.268 | 0.436 | 0.433 |
| 0.019 | 0.027 | 0.067 | 0.166 | 0.27 | 0.358 | 0.367 |
| 0.026 | 0.045 | 0.084 | 0.187 | 0.311 | 0.398 | 0.342 |
| \# Weight Matrix - 2 |  |  |  |  |  |  |
| 0 | 0.0733 | 0.1733 | 0.2992 | 0.446 | 0.5795 | 0.7203 |
| 0 | 0.0785 | 0.1797 | 0.3003 | 0.4468 | 0.5795 | 0.7143 |
| 0 | 0.084 | 0.1873 | 0.311 | 0.4408 | 0.576 | 0.6948 |
| 0 | 0.085 | 0.194 | 0.321 | 0.45 | 0.5813 | 0.6668 |
| 0 | 0.079 | 0.1918 | 0.3163 | 0.4473 | 0.5743 | 0.6628 |
| 0 | 0.0697 | 0.1807 | 0.3038 | 0.4455 | 0.5825 | 0.6998 |
| 0 | 0.0643 | 0.1685 | 0.2907 | 0.4338 | 0.5893 | 0.7485 |
| 0 | 0.0598 | 0.1572 | 0.2857 | 0.4387 | 0.6195 | 0.8123 |
| 0 | 0.0617 | 0.15 | 0.2662 | 0.425 | 0.6262 | 0.8682 |
| 0 | 0.0607 | 0.1497 | 0.2533 | 0.3963 | 0.6057 | 0.8412 |
| 0 | 0.0608 | 0.1473 | 0.24 | 0.355 | 0.5375 | 0.7817 |
| 0 | 0.0545 | 0.1417 | 0.2393 | 0.3318 | 0.4772 | 0.718 |
| 0 | 0.048 | 0.1233 | 0.2218 | 0.3148 | 0.4282 | 0.7055 |
| 0 | 0.0463 | 0.117 | 0.2045 | 0.2927 | 0.3982 | 0.6358 |
| 0 | 0.0462 | 0.118 | 0.2002 | 0.2798 | 0.396 | 0.6755 |
| 0 | 0.0473 | 0.1208 | 0.202 | 0.2695 | 0.3752 | 0.6523 |
| 0 | 0.042 | 0.1142 | 0.205 | 0.2675 | 0.3703 | 0.6678 |
| 0 | 0.0367 | 0.1053 | 0.1952 | 0.258 | 0.3345 | 0.521 |
| 0 | 0.0322 | 0.101 | 0.194 | 0.2598 | 0.3227 | 0.4225 |
| 0 | 0.0313 | 0.0945 | 0.1937 | 0.2632 | 0.3212 | 0.3588 |
| 0 | 0.0312 | 0.0895 | 0.2015 | 0.2742 | 0.3532 | 0.3367 |
| 0 | 0.0293 | 0.0835 | 0.1987 | 0.2888 | 0.3812 | 0.3847 |
| 0 | 0.029 | 0.0992 | 0.2054 | 0.2847 | 0.4021 | 0.4114 |
| 0 | 0.0281 | 0.1007 | 0.2267 | 0.3261 | 0.3847 | 0.4357 |
| 0 | 0.0288 | 0.1045 | 0.2282 | 0.3338 | 0.3984 | 0.4062 |
| 0 | 0.0323 | 0.0918 | 0.2277 | 0.3525 | 0.3862 | 0.3827 |
| 0 | 0.0331 | 0.0939 | 0.2097 | 0.3355 | 0.4296 | 0.5145 |
| 0 | 0.0352 | 0.0901 | 0.2082 | 0.3326 | 0.4961 | 0.7133 |
| 0 | 0.0311 | 0.0815 | 0.2152 | 0.3261 | 0.5283 | 0.9183 |
| 0 | 0.0331 | 0.077 | 0.1989 | 0.3325 | 0.4804 | 0.9181 |
| 0 | 0.0326 | 0.0756 | 0.1883 | 0.3311 | 0.4127 | 0.784 |
| 0 | 0.0313 | 0.078 | 0.175 | 0.3326 | 0.3957 | 0.5933 |
| 0 | 0.032 | 0.0753 | 0.1748 | 0.3127 | 0.3924 | 0.455 |
| 0 | 0.0334 | 0.0808 | 0.1777 | 0.3134 | 0.4162 | 0.4746 |
| 0 | 0.0369 | 0.0836 | 0.1851 | 0.3267 | 0.4009 | 0.5369 |
| 0 | 0.0339 | 0.0805 | 0.1806 | 0.3283 | 0.4403 | 0.6021 |
| 0 | 0.0308 | 0.068 | 0.1713 | 0.3104 | 0.4016 | 0.555 |
| 0 | 0.0306 | 0.0625 | 0.1401 | 0.2712 | 0.3946 | 0.505 |
| 0 | 0.0301 | 0.0658 | 0.133 | 0.2375 | 0.3551 | 0.4336 |
| 0 | 0.029 | 0.0641 | 0.1307 | 0.2187 | 0.3307 | 0.3992 |
| 0 | 0.0252 | 0.062176 | 0.1371 | 0.2185 | 0.3257 | - 0.3888 |
| 0 | 0.0275 | 0.06 | 0.1337 | 0.2251 | 0.3284 | 0.3758 |
| \# Weights-at-Age Pointers |  |  |  |  |  |  |

1
1
1
1
2
2
\# Selectivity Block Assignment
\# Fleet 1 Selectivity Block Assignment
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
2
2
2
2
2
2
2
2
2
2
2
\# Selectivity Block \#1 Data

| 0 | 1 | 0 | 0.25 |
| :--- | :---: | :---: | :---: |
| 0.5 | 1 | 0 | 0.25 |
| 0.9 | 1 | 0 | 0.25 |
| 1 | -1 | 0 | 0.25 |
| 1 | -1 | 0 | 0.25 |
| 1 | -1 | 0 | 0.25 |
| 1 | -1 | 0 | 0.25 |
| 3 | 1 | 0 | 1 |
| 0.5 | 1 | 0 | 1 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |


| 0.2 | 1 | $0 \quad 0.5$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | -1 | 00 |  |  |  |  |  |
| 1 | -1 | 00 |  |  |  |  |  |
| 1 | -1 | 00 |  |  |  |  |  |
| 1 | -1 | 00 |  |  |  |  |  |
| 1 | -1 | 00 |  |  |  |  |  |
| 1 | -1 | 00 |  |  |  |  |  |
| 2 | 1 | $0 \quad 1$ |  |  |  |  |  |
| 0.5 | 1 | $0 \quad 1$ |  |  |  |  |  |
| 0 | 0 | 00 |  |  |  |  |  |
| 0 | 0 | 00 |  |  |  |  |  |
| 0 | 0 | 00 |  |  |  |  |  |
| 0 | 0 | 00 |  |  |  |  |  |
| \# Fleet Start Age |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| \# Fleet End Age |  |  |  |  |  |  |  |
| 7 |  |  |  |  |  |  |  |
| \# Age Range for Average F |  |  |  |  |  |  |  |
| 24 |  |  |  |  |  |  |  |
| \# Average F report option (1=unweighted, 2=Nweighted, 3=Bweighted) |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| \# Use Likelihood constants? (1=yes) |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| \# Release Mortality by Fleet |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| \# Catch Data |  |  |  |  |  |  |  |
| \# Fleet-1 Catch Data |  |  |  |  |  |  |  |
| 0 | 46838 | 28699 | 6533 | 2574 | 350 | 621 | 16737 |
| 9865 | 36138 | 38173 | 16166 | 2622 | 821 | 339 | 21331 |
| 4088 | 13916 | 18658 | 19359 | 6108 | 813 | 400 | 17969 |
| 23847 | 12214 | 11979 | 9109 | 4530 | 1165 | 321 | 12405 |
| 26394 | 52154 | 14250 | 5268 | 2571 | 1045 | 402 | 214999 |
| 12380 | 42001 | 37183 | 8696 | 996 | 675 | 372 | 18169 |
| 28364 | 27417 | 18303 | 10842 | 1877 | 285 | 270 | 12129 |
| 16594 | 50747 | 23438 | 6968 | 1943 | 242 | 111 | 14270 |
| 6922 | 24941 | 27708 | 6989 | 1513 | 396 | 197 | 11856 |
| 17247 | 32346 | 16505 | 13037 | 3645 | 490 | 177 | 713408 |
| 4216 | 41312 | 20957 | 4806 | 1477 | 318 | 128 | 10656 |
| 20451 | 36760 | 22229 | 5975 | 994 | 311 | 84 | 9946 |
| 1497 | 69831 | 32311 | 8056 | 1374 | 462 | 93 | 12791 |
| 12677 | 16721 | 39175 | 7753 | 937 | 135 | 27 | 9230 |
| 3731 | 15308 | 8795 | 20211 | 1830 | 208 | 50 | 7936 |
| 7118 | 18737 | 17110 | 3512 | 4678 | 233 | 21 | 7044 |
| 12861 | 41299 | 14330 | 9408 | 1344 | 1082 | 135 | 7966 |
| 8163 | 26107 | 9591 | 3256 | 1936 | 87 | 79 | 4205 |
| 6096 | 27460 | 5580 | 5277 | 932 | 261 | 95 | 3533 |
| 20852 | 8018 | 4923 | 2835 | 776 | 161 | 121 | 2762 |
| 7321 | 39242 | 5758 | 1566 | 354 | 115 | 25 | 2880 |
| 16940 | 12742 | 4338 | 1108 | 181 | 53 | 20 | 1745 |
| 8538 | 13480 | 2336 | 1710 | 408 | 48 | 42 | 1487 |
| 12389 | 4685 | 1252 | 606 | 199 | 0 | 0 | 996 |
| 19699 | 14938 | 362 | 176 | 150 | 17 | 0 | 1038 |
| 643 | 5797 | 448 | 232 | 98 | 21 | $3 \quad 3$ | 397 |
| 15764 | 20624 | 654 | 109 | 39 | 9 | 1 | 1837 |
| 17436 | 24343 | 787 | 82 | 10 | 3 | 0 | 1547 |
| 10645 | 20032 | 742 | 41 | 5 | 1 | $0 \quad 1$ | 1206 |
| 6622 | 27513 | 1220 | 22 | 4 | 1 | $0 \quad 1$ | 1360 |
| 3946 | 15120 | 2894 | 75 | 4 | 0 | $0 \quad 1$ | 1184 |
| 25982 | 8198 | 675 | 373 | 15 | 3 | 0 | 977 |
| 6637 | 31023 | 819 | 116 | 61 | 12 | 1 | 1399 |
| 8501 | 12026 | 649 | 24 | 5 | 11 | 1981 | 81 |
| 13467 | 27555 | 2450 | 284 | 21 | 1 | 1 | 1974 |
| 3883 | 23595 | 2613 | 267 | 15 | 1 | 1 | 1549 |
| 4509 | 5780 | 4809 | 315 | 25 | 3 | 0 | 780 |
| 3559 | 5871 | 4394 | 290 | 57 | 5 | 1 | 704 |


| 6523 | 7386 | 2571 | 684 | 129 | 29 | 2 | 899 |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| 6429 | 14041 | 4132 | 752 | 129 | 52 | 23 | 1261 |
| 11987 | 26870 | 1036 | 188 | 96 | 18 | 10 | 1118 |
| 11479 | 23168 | 9450 | 1175 | 199 | 58 | 24 | 1662 |

\# Discards
\# Fleet-1 Discards Data

\# Release Proportion \# Fleet-1 Release Data

| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| \# Survey Index Data <br> \# Aggregate Index Units |  |  |  |  |  |  |
| $22222$ |  |  |  |  |  |  |
| \# Age Proportion Index Units |  |  |  |  |  |  |
| $22222$ |  |  |  |  |  |  |
| $11111$ |  |  |  |  |  |  |
| 310595 |  |  |  |  |  |  |
| \# Index Selectivity Link to Fleet |  |  |  |  |  |  |
| \# Index Selectivity Options 1=by age, 2=logisitic, 3=double |  |  |  |  |  |  |
| \# Index Start Age |  |  |  |  |  |  |
| $21112$ |  |  |  |  |  |  |
| \# Index End Age |  |  |  |  |  |  |
| 77127 |  |  |  |  |  |  |
| \# Estimate Proportion (Yes=1) |  |  |  |  |  |  |
| 11000 |  |  |  |  |  |  |
| \# Use Index (Yes=1) |  |  |  |  |  |  |
| 11100 |  |  |  |  |  |  |
| \# Index-1 Selectivity Data |  |  |  |  |  |  |
| 0 | -1 | 0 | 0 |  |  |  |
| 0.5 | 1 | 0 | 0.5 |  |  |  |
| 1 | -1 | 0 | 0 |  |  |  |
| 1 | -1 | 0 | 0 |  |  |  |
| 1 | -1 | 0 | 0 |  |  |  |
| 1 | -1 | 0 | 0 |  |  |  |
| 1 | -1 | 0 | 0 |  |  |  |
| 2 | 1 | 0 | 1 |  |  |  |
| 0.5 | 1 | 0 | 1 |  |  |  |
| 0 | 0 | 0 | 0 |  |  |  |
| 0 | 0 | 0 | 0 |  |  |  |
| 0 | 0 | 0 | 0 |  |  |  |
| 0 | 0 | 0 | 0 |  |  |  |
| \# Index-2 Selectivity Data |  |  |  |  |  |  |
| 0.2 | 1 | 0 | 0.5 |  |  |  |
| 0.5 | 1 | 0 | 0.5 |  |  |  |
| 1 | -1 | 0 | 0 |  |  |  |
| 1 | -1 | 0 | 0 |  |  |  |


| 1 | -1 | 0 | 0 |
| :--- | :---: | :--- | :--- |
| 1 | -1 | 0 | 0 |
| 1 | -1 | 0 | 0 |
| 2 | 1 | 0 | 1 |
| 0.5 | 1 | 0 | 1 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |

\# Index-3 Selectivity Data

| 1 | -1 | 0 | 0 |
| :--- | :--- | :--- | :--- |
| 0 | -1 | 0 | 0 |
| 0 | -1 | 0 | 0 |
| 0 | -1 | 0 | 0 |
| 0 | -1 | 0 | 0 |
| 0 | -1 | 0 | 0 |
| 0 | -1 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |


| \# Index-4 Selectivity Data |  |  |  |
| :--- | :--- | :--- | :--- |
| 1 | -1 | 0 | 1 |
| 1 | -1 | 0 | 0 |
| 1 | -1 | 0 | 0 |
| 1 | -1 | 0 | 0 |
| 1 | -1 | 0 | 0 |
| 1 | -1 | 0 | 0 |
| 1 | -1 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |

\# Index-5 Selectivity Data

| 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 4 | 1 | 0 | 1 |
| 0.5 | 1 | 0 | 1 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |

\# Index-1 Data

| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1984 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1985 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1986 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1991 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1993 | 1474 | 0.3 | 0 | 0.452 | 0.482 | 0.055 | 0.008 | 0.003 | 0.001 | 50 |  |


| 19942431 | 0.3 | 0 | 0.742 | 0.108 | 0.123 | 0.018 | 0.005 | 0.003 | 50 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19952615 | 0.3 | 0 | 0.646 | 0.243 | 0.067 | 0.034 | 0.008 | 0.002 | 50 |
| 19962093 | 0.3 | 0 | 0.702 | 0.16 | 0.102 | 0.017 | 0.018 | 0.003 | 50 |
| 19973169 | 0.3 | 0 | 0.444 | 0.485 | 0.049 | 0.017 | 0.001 | 0.004 | 50 |
| 19982694 | 0.3 | 0 | 0.551 | 0.28 | 0.154 | 0.011 | 0.003 | 0.001 | 50 |
| 19991900 | 0.3 | 0 | 0.721 | 0.196 | 0.059 | 0.022 | 0.002 | 0.001 | 50 |
| 20002925 | 0.3 | 0 | 0.787 | 0.14 | 0.062 | 0.009 | 0.001 | 0 | 50 |
| 20011909 | 0.3 | 0 | 0.558 | 0.365 | 0.065 | 0.007 | 0.003 | 0.001 | 50 |
| 20023235 | 0.3 | 0 | 0.713 | 0.212 | 0.054 | 0.016 | 0.003 | 0 | 50 |
| 20032543 | 0.3 | 0 | 0.588 | 0.356 | 0.051 | 0.004 | 0.001 | 0 | 50 |
| 20041907 | 0.3 | 0 | 0.844 | 0.121 | 0.032 | 0.001 | 0.001 | 0 | 50 |
| 2005857 | 0.3 | 0 | 0.804 | 0.145 | 0.033 | 0.014 | 0.003 | 0 | 50 |
| 20061232 | 0.3 | 0 | 0.779 | 0.191 | 0.025 | 0.005 | 0 | 0 | 50 |
| 20071081 | 0.3 | 0 | 0.837 | 0.147 | 0.014 | 0.002 | 0 | 0 | 50 |
| 20081152 | 0.3 | 0 | 0.657 | 0.301 | 0.039 | 0.002 | 0 | 0 | 50 |
| 20091382 | 0.3 | 0 | 0.769 | 0.203 | 0.026 | 0.001 | 0 | 0 | 50 |
| 20101348 | 0.3 | 0 | 0.549 | 0.405 | 0.038 | 0.003 | 0.005 | 0 | 50 |
| 2011783 | 0.3 | 0 | 0.749 | 0.2 | 0.046 | 0.005 | 0.001 | 0 | 50 |
| 20121376 | 0.3 | 0 | 0.707 | 0.258 | 0.031 | 0.004 | 0.001 | 0 | 50 |
| 20131350 | 0.3 | 0 | 0.466 | 0.481 | 0.049 | 0.003 | 0 | 0 | 50 |
| 20141361 | 0.3 | 0 | 0.677 | 0.27 | 0.049 | 0.003 | 0 | 0 | 50 |
| 20153288 | 0.3 | 0 | 0.851 | 0.143 | 0.006 | 0.001 | 0 | 0 | 50 |
| 20162374 | 0.3 | 0 | 0.594 | 0.39 | 1E-06 | 0.016 | 0.001 | 0 | 50 |
| 20171874 | 0.3 | 0 | 0.474 | 0.444 | 0.076 | 0.006 | 0.001 | 0 | 50 |
| 2018868 | 0.3 | 0 | 0.497 | 0.342 | 0.138 | 0.021 | 0.003 | 0 | 50 |
| 20191797 | 0.3 | 0 | 0.316 | 0.463 | 0.193 | 0.024 | 0.003 | 0 | 50 |
| 20202294 | 0.3 | 0 | 0.686 | 0.254 | 0.055 | 0.004 | 0 | 0 | 50 |
| 20211620 | 0.3 | 0 | 0.351 | 0.587 | 0.053 | 0.006 | 0.002 | 0.002 | 50 |
| \# Index-2 Da |  |  |  |  |  |  |  |  |  |
| 19800 | 0 | 0 | $0 \quad 0$ | 0 | 0 | 0 | 0 | 0 |  |
| 19810 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 |  |
| 19820 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 |  |
| 19830 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 |  |
| 19840 | 0 | 0 | $0 \quad 0$ | 0 | 0 | 0 | 0 | 0 |  |
| 19850 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 |  |
| 19860 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 |  |
| 19870 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 |  |
| 19880 | 0 | 0 | $0 \quad 0$ | 0 | 0 | 0 | 0 | 0 |  |
| 19890 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 |  |
| 19900 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 |  |
| 19910 | 0 | 0 | $0 \quad 0$ | 0 | 0 | 0 | 0 | 0 |  |
| 19920 | 0 | 0 | $0 \quad 0$ | 0 | 0 | 0 | 0 | 0 |  |
| 19932309 | 0.3 | 0.309 | 0.451 | 0.206 | 0.029 | 0.004 | 0.001 | 0 | 50 |
| 19942829 | 0.3 | 0.393 | 0.467 | 0.074 | 0.053 | 0.012 | 0.001 | 0 | 50 |
| 19953841 | 0.3 | 0.813 | 0.124 | 0.043 | 0.008 | 0.009 | 0.001 | 0 | 50 |
| 19963103 | 0.3 | 0.743 | 0.191 | 0.043 | 0.017 | 0.003 | 0.002 | 0 | 50 |
| 19973889 | 0.3 | 0.675 | 0.174 | 0.128 | 0.016 | 0.005 | 0.001 | 0.00 | 150 |
| 19983566 | 0.3 | 0.803 | 0.131 | 0.043 | 0.02 | 0.002 | 0.001 | 0 | 50 |
| 19993851 | 0.3 | 0.644 | 0.28 | 0.05 | 0.013 | 0.011 | 0.001 | 0 | 50 |
| 20003631 | 0.3 | 0.654 | 0.299 | 0.035 | 0.006 | 0.005 | 0.002 | 0.00 | 150 |
| 20017322 | 0.3 | 0.868 | 0.09 | 0.037 | 0.004 | 0.001 | 0 | 0 | 50 |
| 20024331 | 0.3 | 0.622 | 0.305 | 0.062 | 0.01 | 0.001 | 0 | 0 | 50 |
| 20037017 | 0.3 | 0.631 | 0.224 | 0.131 | 0.011 | 0.002 | 0 | 0 | 50 |
| 20045545 | 0.3 | 0.804 | 0.126 | 0.048 | 0.021 | 0.001 | 0 | 0 | 50 |
| 20053066 | 0.3 | 0.775 | 0.159 | 0.06 | 0.005 | 0 | 0 | 0 | 50 |
| 20063050 | 0.3 | 0.934 | 0.047 | 0.015 | 0.003 | 0.001 | 0 | 0 | 50 |
| 20073296 | 0.3 | 0.656 | 0.291 | 0.045 | 0.005 | 0.001 | 0.001 | 0 | 50 |
| 20086315 | 0.3 | 0.773 | 0.208 | 0.018 | 0.001 | 0 | 0 | 0 | 50 |
| 20092838 | 0.3 | 0.791 | 0.18 | 0.025 | 0.003 | 0.001 | 0 | 0 | 50 |
| 20102855 | 0.3 | 0.797 | 0.109 | 0.091 | 0.003 | 0 | 0 | 0 | 50 |
| 20114053 | 0.3 | 0.872 | 0.086 | 0.034 | 0.006 | 0.001 | 0 | 0 | 50 |
| 20121894 | 0.3 | 0.703 | 0.213 | 0.071 | 0.01 | 0.003 | 0 | 0 | 50 |
| 20137680 | 0.3 | 0.929 | 0.046 | 0.02 | 0.004 | 0 | 0 | 0 | 50 |
| 20145157 | 0.3 | 0.873 | 0.098 | 0.026 | 0.002 | 0 | 0 | 0 | 50 |
| 20153816 | 0.3 | 0.734 | 0.234 | 0.03 | 0.002 | 0 | 0 | 0 | 50 |
| 20163817 | 0.3 | 0.712 | 0.225 | 0.053 | 0.008 | 0 | 0 | 0 | 50 |



| 1997409 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998893 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999550 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000320 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001585 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002280 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003456 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004917 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005849 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20061010 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007339 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008780 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009389 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010324 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20111002 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012442 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20131535 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2014261 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015211 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016666 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017489 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018661.75 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019307 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20200 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021340 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| \# Index-5 Data |  |  |  |  |  |  |  |  |  |
| 19800 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19810 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19820 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19830 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19840 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19850 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19860 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19870 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19880 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19890 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19900 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19910 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19920 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19930 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19940 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19950 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19960 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19970 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20020 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20030 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20040 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 200538.66 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 200672.953 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 200721.87 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 200825.23 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009164.82 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010111.12 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 201148.6 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012139.25 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 201382.85 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20140 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20150 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20160 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20170 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20180 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20190 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

```
20200 0
20210
# Phase Control
# Phase for F mult in 1st Year
1
# Phase for F mult Deviations
1
# Phase for Recruitment Deviations
3
# Phase for N in 1st Year
1
# Phase for Catchability in 1st Year
2
# Phase for Catchability Deviations
-5
# Phase for Stock-Recruitment Relationship
3
# Phase for Steepness
-5
# Recruitment CV by Year
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
1
# Lambdas by Index
11111
# Lambda for Total Catch in Weight by Fleet
1
```



|  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |



```
# CV for Deviation from Initial Steepness
. }
# Lambda for Deviation from Unexploited Stock Size
O
# CV for Deviation from Unexploited Stock Size
.9
# NAA Deviations Flag
1
# Initial Numbers at Age in 1st Year
1 0 0 0 0 0 0 5 0 0 0 0 0 2 5 0 0 0 0 1 2 5 0 0 0 6 0 0 0 0 3 0 0 0 0 1 0 0 0 0
# Initial F Mult in 1st Year by Fleet
1
# Initial Catchabilty by Index
.001 .001 .001 .001 0.001
# Stock-Recruitment Flag
O
# Initial Unexploited Stock
1000
# Initial Steepness
1
# Maximum F
2.5
# Ignore Guesses (Yes=1)
O
# Projection Control
# Do Projections (Yes=1)
O
# Fleet Directed Flag
1
# Final Year in Projection
2022
# Projection Data by Year
2022
# Do MCMC (Yes=1)
O
# MCMC Year Option
1
# MCMC Iterations
O
# MCMC Thinning Factor
O
# MCMC Random Seed
O
# Agepro R Option
-1
# Agepro R Option Start Year
O
# Agepro R Option End Year
O
# Export R Flag
1
# Test Value
-23456
######
###### FINIS ######
# Fleet Names
#$All
# Survey Names
#$NI-Q1
#$NI_Q2
#$NI-MIK
#$UK-BTS
#$UK-FSP
#
```

Table 37.7. Whiting 7.a. Selectivity of the catches and indices.

| Age | Catch | NI-Q1 | NI-Q4 | NI-MIK |
| :--- | :--- | :--- | :--- | :--- |
| 0 | 0.115 | 0.000 | 0.669 | 1.000 |
| 1 | 0.841 | 0.478 | 0.760 | 0.000 |
| 2 | 1.995 | 1.000 | 1.000 | 1.000 |
| 4 | 1.000 | 1.000 | 1.000 | 0.000 |
| 5 | 1.000 | 1.000 | 1.000 | 0.000 |
| 6 |  |  |  | 0.000 |

Table 37.8. Whiting 7.a Fishing mortality- (F) -at age.

|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.024 | 0.17 | 0.409 | 0.472 | 0.478 | 0.479 | 0.479 |
| 1981 | 0.029 | 0.207 | 0.497 | 0.573 | 0.582 | 0.583 | 0.583 |
| 1982 | 0.033 | 0.239 | 0.575 | 0.663 | 0.673 | 0.674 | 0.674 |
| 1983 | 0.034 | 0.248 | 0.596 | 0.687 | 0.697 | 0.698 | 0.698 |
| 1984 | 0.041 | 0.297 | 0.715 | 0.825 | 0.837 | 0.838 | 0.838 |
| 1985 | 0.05 | 0.359 | 0.865 | 0.997 | 1.011 | 1.013 | 1.013 |
| 1986 | 0.039 | 0.281 | 0.677 | 0.78 | 0.791 | 0.792 | 0.793 |
| 1987 | 0.041 | 0.296 | 0.712 | 0.821 | 0.833 | 0.834 | 0.834 |
| 1988 | 0.035 | 0.252 | 0.607 | 0.7 | 0.71 | 0.711 | 0.711 |
| 1989 | 0.049 | 0.355 | 0.854 | 0.985 | 0.999 | 1.001 | 1.001 |
| 1990 | 0.042 | 0.303 | 0.729 | 0.841 | 0.853 | 0.854 | 0.854 |
| 1991 | 0.043 | 0.307 | 0.74 | 0.853 | 0.866 | 0.867 | 0.867 |
| 1992 | 0.07 | 0.506 | 1.217 | 1.403 | 1.424 | 1.426 | 1.426 |
| 1993 | 0.056 | 0.403 | 0.97 | 1.118 | 1.134 | 1.136 | 1.136 |
| 1994 | 0.058 | 0.418 | 1.007 | 1.161 | 1.178 | 1.179 | 1.179 |
| 1995 | 0.103 | 0.758 | 0.898 | 0.902 | 0.902 | 0.902 | 0.902 |
| 1996 | 0.112 | 0.824 | 0.975 | 0.979 | 0.979 | 0.979 | 0.979 |
| 1997 | 0.098 | 0.717 | 0.849 | 0.853 | 0.853 | 0.853 | 0.853 |
| 1998 | 0.138 | 1.01 | 1.195 | 1.201 | 1.201 | 1.201 | 1.201 |
| 1999 | 0.109 | 0.797 | 0.943 | 0.948 | 0.948 | 0.948 | 0.948 |
| 2000 | 0.141 | 1.036 | 1.226 | 1.232 | 1.232 | 1.232 | 1.232 |
| 2001 | 0.115 | 0.841 | 0.995 | 0.999 | 1 | 1 | 1 |
| 2002 | 0.157 | 1.151 | 1.362 | 1.368 | 1.368 | 1.368 | 1.368 |
| 2003 | 0.078 | 0.575 | 0.681 | 0.684 | 0.684 | 0.684 | 0.684 |
| 2004 | 0.22 | 1.612 | 1.908 | 1.917 | 1.917 | 1.917 | 1.917 |
| 2005 | 0.06 | 0.437 | 0.517 | 0.519 | 0.519 | 0.519 | 0.519 |
| 2006 | 0.193 | 1.418 | 1.679 | 1.686 | 1.686 | 1.686 | 1.686 |
| 2007 | 0.152 | 1.118 | 1.323 | 1.329 | 1.329 | 1.329 | 1.329 |
| 2008 | 0.124 | 0.907 | 1.073 | 1.078 | 1.078 | 1.078 | 1.078 |


|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2009 | 0.14 | 1.026 | 1.215 | 1.22 | 1.22 | 1.22 | 1.22 |
| 2010 | 0.146 | 1.074 | 1.271 | 1.277 | 1.277 | 1.277 | 1.277 |
| 2011 | 0.11 | 0.805 | 0.952 | 0.957 | 0.957 | 0.957 | 0.957 |
| 2012 | 0.145 | 1.063 | 1.258 | 1.263 | 1.264 | 1.264 | 1.264 |
| 2013 | 0.084 | 0.613 | 0.726 | 0.729 | 0.729 | 0.729 | 0.729 |
| 2015 | 0.124 | 0.913 | 1.401 | 1.658 | 1.666 | 1.666 | 1.666 |

Table 37.9. Whiting 7.a Stock Numbers-at-age (start of year) ('1000).

|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 543303 | 406806 | 132581 | 22541 | 7971 | 1082 | 1911 |
| 1981 | 317076 | 180565 | 153764 | 42954 | 7658 | 2839 | 1104 |
| 1982 | 281004 | 104844 | 65792 | 45607 | 13181 | 2460 | 1312 |
| 1983 | 881826 | 92499 | 36982 | 18048 | 12790 | 3864 | 1145 |
| 1984 | 632618 | 289933 | 32352 | 9940 | 4944 | 3661 | 1485 |
| 1985 | 513219 | 206571 | 96500 | 7717 | 2373 | 1231 | 1326 |
| 1986 | 872068 | 166145 | 64613 | 19821 | 1550 | 496 | 553 |
| 1987 | 472746 | 285395 | 56194 | 16020 | 4947 | 404 | 283 |
| 1988 | 484038 | 154396 | 95116 | 13447 | 3838 | 1236 | 178 |
| 1989 | 593233 | 159045 | 53755 | 25286 | 3637 | 1085 | 414 |
| 1990 | 516210 | 192163 | 49963 | 11157 | 5141 | 769 | 328 |
| 1991 | 665548 | 168424 | 63587 | 11752 | 2620 | 1259 | 278 |
| 1992 | 230445 | 217013 | 55482 | 14796 | 2726 | 634 | 385 |
| 1993 | 212078 | 73102 | 58637 | 8012 | 1980 | 377 | 146 |
| 1994 | 183038 | 68243 | 21891 | 10846 | 1426 | 366 | 100 |
| 1995 | 339009 | 58772 | 20123 | 3901 | 1850 | 252 | 85 |
| 1996 | 202658 | 104021 | 12334 | 4000 | 862 | 431 | 82 |
| 1997 | 171146 | 61633 | 20452 | 2270 | 818 | 186 | 115 |
| 1998 | 167120 | 52808 | 13475 | 4267 | 527 | 200 | 76 |
| 1999 | 208186 | 49551 | 8619 | 1989 | 699 | 91 | 50 |
| 2000 | 109678 | 63544 | 10005 | 1637 | 420 | 156 | 32 |
| 2001 | 192438 | 32402 | 10101 | 1431 | 260 | 70 | 33 |
| 2002 | 79684 | 58389 | 6263 | 1822 | 287 | 55 | 23 |
| 2003 | 121206 | 23176 | 8277 | 782 | 252 | 42 | 12 |
| 2004 | 94124 | 38131 | 5841 | 2043 | 215 | 73 | 16 |
| 2005 | 106078 | 25706 | 3407 | 423 | 164 | 18 | 8 |
| 2006 | 154477 | 34009 | 7441 | 991 | 137 | 56 | 9 |
| 2007 | 103890 | 43320 | 3689 | 677 | 100 | 15 | 7 |
| 2008 | 148979 | 30352 | 6345 | 479 | 98 | 15 | 3 |


|  | Age 0 | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 92763 | 44796 | 5490 | 1058 | 89 | 19 | 4 |
| 2010 | 92375 | 27442 | 7191 | 795 | 170 | 15 | 4 |
| 2011 | 151247 | 27150 | 4200 | 984 | 121 | 27 | 3 |
| 2012 | 78928 | 46116 | 5440 | 790 | 206 | 27 | 7 |
| 2013 | 161013 | 23234 | 7139 | 754 | 122 | 33 | 6 |
| 2014 | 196488 | 50394 | 5638 | 1685 | 198 | 34 | 11 |
| 2015 | 123220 | 55230 | 5561 | 524 | 173 | 22 | 5 |
| 2016 | 81525 | 37021 | 9933 | 921 | 96 | 34 | 5 |
| 2017 | 103030 | 25846 | 9872 | 2621 | 271 | 30 | 13 |
| 2018 | 107633 | 33131 | 7649 | 2947 | 872 | 95 | 15 |
| 2019 | 214378 | 34613 | 9808 | 2285 | 981 | 306 | 40 |
| 2020 | 107078 | 66471 | 7843 | 2135 | 553 | 251 | 92 |
| 2021 | 196382 | 33260 | 15257 | 1733 | 525 | 143 | 92 |
| 2022 | 119242 | 60507 | 7194 | 3143 | 397 | 127 | 59 |

Table 37.10. Whiting 7.a Stock Summary: weights in tonnes: CatchPred is predicted catch from ASAP. Recruitment-at-age zero ('1000), $\mathrm{F}_{\text {bar }}$ ages (1-3).

| Year | Lan | Dis | Cat | CatPred | Tsb | Ssb | SsbCv | Recr | RecrCv | Fbar | FbarCv |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 13422 | 3314 | 16737 | 16777.68 | 65098.12 | 35279.22 | 0.318856 | 543303.1 | 0.352142 | 0.350178 | 0.319952 |
| 1981 | 18267 | 3064 | 21331 | 21260.44 | 60560.46 | 46386.09 | 0.240546 | 317076.1 | 0.425481 | 0.425772 | 0.292762 |
| 1982 | 17167 | 801 | 17969 | 17869.68 | 43451.88 | 34644.96 | 0.273053 | 281003.5 | 0.446187 | 0.492646 | 0.332003 |
| 1983 | 10577 | 1829 | 12405 | 12315.84 | 29595.21 | 21732.78 | 0.352274 | 881825.6 | 0.25352 | 0.510116 | 0.391793 |
| 1984 | 11619 | 3380 | 14999 | 14698.42 | 37551.51 | 14646.8 | 0.421607 | 632618.3 | 0.314123 | 0.612299 | 0.353879 |
| 1985 | 15525 | 2644 | 18169 | 17939.59 | 36881.93 | 22483.94 | 0.297368 | 513218.8 | 0.344648 | 0.74031 | 0.324985 |
| 1986 | 10063 | 2066 | 12129 | 12067.14 | 28711.32 | 18028.21 | 0.325739 | 872068.1 | 0.262285 | 0.579194 | 0.350435 |
| 1987 | 10411 | 3859 | 14270 | 14058.48 | 33127.54 | 16060.94 | 0.341842 | 472746.4 | 0.347163 | 0.609553 | 0.323548 |
| 1988 | 10245 | 1611 | 11856 | 11796.65 | 29932.84 | 20406.6 | 0.281277 | 484038.1 | 0.323115 | 0.519509 | 0.326943 |
| 1989 | 11305 | 2103 | 13408 | 13392.93 | 26552.4 | 16898.37 | 0.309355 | 593233.1 | 0.267113 | 0.731426 | 0.32207 |
| 1990 | 8212 | 2444 | 10656 | 10633.81 | 24215.74 | 12532.24 | 0.330763 | 516210.1 | 0.250499 | 0.624308 | 0.305843 |
| 1991 | 7348 | 2598 | 9946 | 9911.575 | 22671.57 | 13492.49 | 0.263406 | 665548.1 | 0.168403 | 0.633543 | 0.252974 |
| 1992 | 8588 | 4203 | 12791 | 12555.59 | 21940.34 | 11523.71 | 0.210731 | 230444.8 | 0.151741 | 1.041951 | 0.182648 |
| 1993 | 6523 | 2707 | 9230 | 6749.623 | 12706.2 | 9321.555 | 0.151584 | 212078.4 | 0.130277 | 0.83005 | 0.166388 |
| 1994 | 6763 | 1173 | 7936 | 4983.258 | 8518.981 | 5366.164 | 0.159371 | 183037.6 | 0.136398 | 0.861892 | 0.173015 |
| 1995 | 4893 | 2151 | 7044 | 4536.276 | 6647.761 | 3867.851 | 0.16208 | 339008.6 | 0.117761 | 0.852515 | 0.168255 |
| 1996 | 4335 | 3631 | 7966 | 4376.071 | 7042.328 | 2673.437 | 0.185084 | 202658.5 | 0.128018 | 0.925821 | 0.146303 |
| 1997 | 2277 | 1928 | 4205 | 3001.332 | 5191.624 | 2929.691 | 0.150122 | 171146 | 0.13877 | 0.806479 | 0.164046 |
| 1998 | 2229 | 1304 | 3533 | 2885.702 | 4122.945 | 2422.512 | 0.158984 | 167119.8 | 0.129697 | 1.135113 | 0.16357 |
| 1999 | 1670 | 1092 | 2762 | 2246.251 | 2981.87 | 1430.936 | 0.196543 | 208186.5 | 0.127266 | 0.895955 | 0.175087 |
| 2000 | 762 | 2118 | 2880 | 2354.155 | 3388.852 | 1406.278 | 0.189066 | 109678 | 0.13846 | 1.164767 | 0.150657 |
| 2001 | 733 | 1012 | 1745 | 1619.083 | 2191.775 | 1242.383 | 0.171936 | 192438.2 | 0.130016 | 0.944997 | 0.180667 |
| 2002 | 747 | 740 | 1487 | 1888.927 | 2801.794 | 1108.518 | 0.193556 | 79683.62 | 0.135574 | 1.293595 | 0.154592 |
| 2003 | 517 | 480 | 996 | 1278.184 | 1765.711 | 1114.463 | 0.170001 | 121206.1 | 0.148153 | 0.646684 | 0.234411 |
| 2004 | 133 | 905 | 1038 | 2065.423 | 2282.334 | 1184.153 | 0.196199 | 94124.07 | 0.123805 | 1.812313 | 0.161633 |
| 2005 | 125 | 272 | 397 | 508.3343 | 1307.047 | 476.7326 | 0.256412 | 106078.1 | 0.137889 | 0.490919 | 0.275015 |
| 2006 | 64 | 1773 | 1837 | 2567.899 | 2106.984 | 981.298 | 0.215694 | 154476.8 | 0.125462 | 1.594417 | 0.177927 |
| 2007 | 35 | 1512 | 1547 | 1521.477 | 2043.839 | 518.9917 | 0.270544 | 103889.6 | 0.130456 | 1.256744 | 0.165539 |


| Year | Lan | Dis | Cat | CatPred | Tsb | Ssb | SsbCv | Recr | RecrCv | Fbar | FbarCv |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 37 | 1169 | 1206 | 1284.081 | 1607.203 | 663.2622 | 0.206766 | 148978.8 | 0.124407 | 1.019516 | 0.189031 |
| 2009 | 39 | 1321 | 1360 | 1496.776 | 2158.052 | 675.3145 | 0.212731 | 92762.62 | 0.133211 | 1.153808 | 0.171502 |
| 2010 | 30 | 1154 | 1184 | 1450.439 | 1653.5 | 758.9044 | 0.19121 | 92374.78 | 0.127546 | 1.207466 | 0.184154 |
| 2011 | 31 | 946 | 977 | 979.999 | 1402.276 | 552.4963 | 0.22963 | 151247 | 0.125338 | 0.904552 | 0.193787 |
| 2012 | 60 | 1339 | 1399 | 1433.416 | 2101.441 | 625.7129 | 0.211854 | 78928.24 | 0.134502 | 1.194523 | 0.167598 |
| 2013 | 33 | 948 | 981 | 1036.867 | 1541.642 | 765.628 | 0.196989 | 161012.6 | 0.139679 | 0.68917 | 0.219032 |
| 2014 | 23 | 1951 | 1974 | 2639.152 | 2727.158 | 867.623 | 0.21061 | 196488.1 | 0.117564 | 1.575067 | 0.172063 |
| 2015 | 28 | 1521 | 1549 | 1736.563 | 2484.049 | 611.7477 | 0.251313 | 123219.9 | 0.115493 | 1.026047 | 0.175304 |
| 2016 | 15 | 765 | 780 | 818.2394 | 2019.811 | 879.5625 | 0.179828 | 81525.18 | 0.12617 | 0.583254 | 0.200581 |
| 2017 | 36 | 668 | 704 | 697.6882 | 1866.634 | 1075.751 | 0.162919 | 103030.4 | 0.122207 | 0.466113 | 0.212453 |
| 2018 | 46 | 853 | 899 | 876.8681 | 2139.994 | 1142.747 | 0.167832 | 107632.6 | 0.138471 | 0.465702 | 0.210078 |
| 2019 | 172 | 1089 | 1261 | 1419.745 | 2262.81 | 1259.039 | 0.16592 | 214378.1 | 0.135569 | 0.766307 | 0.226476 |
| 2020 | 88 | 1030 | 1118 | 1223.628 | 2693.506 | 1018.429 | 0.220133 | 107077.7 | 0.173239 | 0.751817 | 0.265065 |
| 2021 | 91 | 1571 | 1662 | 1674.597 | 2261.635 | 1346.999 | 0.209926 | 196381.9 | 0.237435 | 0.818499 | 0.295761 |
| 2022* | NA | NA | NA | NA | NA | 991.7084 | NA | 119242 | NA | 0.778874 | NA |

Table 37.11. Whiting 7.a. Input values for short-term forecast. Note that Sel and CWt refer to the landings and DSel and DCWt refer to the discards. Numbers in thousands; Weights in kg.

| 2022 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt | DSel | DCWt |
| 0 | 119242 | 1.078 | 0 | 0 | 0 | 0 | 0 | 0 | 0.094 | 0.022 |
| 1 | 60507 | 0.803 | 0 | 0 | 0 | 0.028 | 0.072 | 0.062 | 0.621 | 0.035 |
| 2 | 7194 | 0.718 | 1 | 0 | 0 | 0.059 | 0.342 | 0.256 | 0.478 | 0.084 |
| 3 | 3143 | 0.608 | 1 | 0 | 0 | 0.132 | 0.598 | 0.335 | 0.226 | 0.134 |
| 4 | 397 | 0.554 | 1 | 0 | 0 | 0.225 | 0.751 | 0.358 | 0.073 | 0.178 |
| 5 | 127 | 0.518 | 1 | 0 | 0 | 0.328 | 0.807 | 0.43 | 0.017 | 0.146 |
| 6 | 59 | 0.518 | 1 | 0 | 0 | 0.384 | 0.814 | 0.381 | 0.009 | 0 |
| 2023 |  |  |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt | DSel | DCWt |
| 0 | 119242 | 1.078 | 0 | 0 | 0 | 0 | 0 | 0 | 0.094 | 0.022 |
| 1 | 36917 | 0.803 | 0 | 0 | 0 | 0.028 | 0.072 | 0.062 | 0.621 | 0.035 |
| 2 | 13558 | 0.718 | 1 | 0 | 0 | 0.059 | 0.342 | 0.256 | 0.478 | 0.084 |
| 3 | 1545 | 0.608 | 1 | 0 | 0 | 0.132 | 0.598 | 0.335 | 0.226 | 0.134 |
| 4 | 751 | 0.554 | 1 | 0 | 0 | 0.225 | 0.751 | 0.358 | 0.073 | 0.178 |
| 5 | 100 | 0.518 | 1 | 0 | 0 | 0.328 | 0.807 | 0.43 | 0.017 | 0.146 |
| 6 | 49 | 0.518 | 1 | 0 | 0 | 0.384 | 0.814 | 0.381 | 0.009 | 0 |
| 2024 |  |  |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt | DSel | DCWt |
| 0 | 119242 | 1.078 | 0 | 0 | 0 | 0 | 0 | 0 | 0.094 | 0.022 |
| 1 | 36917 | 0.803 | 0 | 0 | 0 | 0.028 | 0.072 | 0.062 | 0.621 | 0.035 |
| 2 | 8272 | 0.718 | 1 | 0 | 0 | 0.059 | 0.342 | 0.256 | 0.478 | 0.084 |
| 3 | 2912 | 0.608 | 1 | 0 | 0 | 0.132 | 0.598 | 0.335 | 0.226 | 0.134 |
| 4 | 369 | 0.554 | 1 | 0 | 0 | 0.225 | 0.751 | 0.358 | 0.073 | 0.178 |
| 5 | 189 | 0.518 | 1 | 0 | 0 | 0.328 | 0.807 | 0.43 | 0.017 | 0.146 |
| 6 | 39 | 0.518 | 1 | 0 | 0 | 0.384 | 0.814 | 0.381 | 0.009 | 0 |

Table 37.12 Whiting 7.a. Single-option output of the short-term forecast ( $F=$ mean $F$ 2019-2021). Numbers in thousands, weights in tonnes.

| 2022 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | F (lan) | CatchNos | Yield | DF | DCatchNos | DYield | StockNos | Biomass | SSNos | SSB |
| 0 | 0 | 0 | 0 | 0.094 | 6635 | 144 | 119242 | 0 | 0 | 0 |
| 1 | 0.072 | 0 | 0 | 0.621 | 21745 | 761 | 60507 | 1700 | 0 | 0 |
| 2 | 0.342 | 97 | 25 | 0.478 | 2915 | 244 | 7194 | 424 | 7194 | 424 |
| 3 | 0.598 | 553 | 186 | 0.226 | 823 | 110 | 3143 | 414 | 3143 | 414 |
| 4 | 0.751 | 136 | 49 | 0.073 | 41 | 7 | 397 | 89 | 397 | 89 |
| 5 | 0.807 | 53 | 23 | 0.017 | 4 | 1 | 127 | 42 | 127 | 42 |
| 6 | 0.814 | 27 | 10 | 0.009 | 0 | 0 | 59 | 23 | 59 | 23 |
| Total | 0.337 | 866 | 293 | 0.442 | 32163 | 1267 | 190669 | 2692 | 10920 | 992 |
| 2023 |  |  |  |  |  |  |  |  |  |  |
| Age | F (lan) | CatchNos | Yield | DF | DCatchNos | DYield | StockNos | Biomass | SSNos | SSB |
| 0 | 0 | 0 | 0 | 0.094 | 6635 | 144 | 119242 | 0 | 0 | 0 |
| 1 | 0.072 | 0 | 0 | 0.621 | 13267 | 464 | 36917 | 1037 | 0 | 0 |
| 2 | 0.342 | 183 | 47 | 0.478 | 5493 | 460 | 13558 | 799 | 13558 | 799 |
| 3 | 0.598 | 272 | 91 | 0.226 | 405 | 54 | 1545 | 204 | 1545 | 204 |
| 4 | 0.751 | 258 | 92 | 0.073 | 78 | 14 | 751 | 169 | 751 | 169 |
| 5 | 0.807 | 42 | 18 | 0.017 | 3 | 0 | 100 | 33 | 100 | 33 |
| 6 | 0.814 | 22 | 8 | 0.009 | 0 | 0 | 49 | 19 | 49 | 19 |
| Total | 0.337 | 777 | 256 | 0.442 | 25881 | 1136 | 172162 | 2261 | 16003 | 1224 |
| 2024 |  |  |  |  |  |  |  |  |  |  |
| Age | F (lan) | CatchNos | Yield | DF | DCatchNos | DYield | StockNos | Biomass | SSNos | SSB |
| 0 | 0 | 0 | 0 | 0.094 | 6635 | 144 | 119242 | 0 | 0 | 0 |
| 1 | 0.072 | 0 | 0 | 0.621 | 13267 | 464 | 36917 | 1037 | 0 | 0 |
| 2 | 0.342 | 111 | 28 | 0.478 | 3352 | 280 | 8272 | 488 | 8272 | 488 |
| 3 | 0.598 | 513 | 172 | 0.226 | 763 | 102 | 2912 | 384 | 2912 | 384 |
| 4 | 0.751 | 127 | 45 | 0.073 | 38 | 7 | 369 | 83 | 369 | 83 |
| 5 | 0.807 | 80 | 34 | 0.017 | 6 | 1 | 189 | 62 | 189 | 62 |
| 6 | 0.814 | 18 | 7 | 0.009 | 0 | 0 | 39 | 15 | 39 | 15 |
| Total | 0.337 | 849 | 286 | 0.442 | 24061 | 998 | 167940 | 2069 | 11781 | 1032 |

Table 37.13. Whiting 7.a. Management options table. Weights in tonnes.

| Fmult | Catch23 | Land23 | Dis23 | Basis | FCatch23 | FLand23 | FDis23 | SSB24 | dSSB | dTac |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 |  | 0 | NA | NA | 2211 | 80.79\% | -100\% |
| 0.1 | 181 | 35 | 146 |  | 0.078 | 0.0119 | 0.066 | 2048 | 67.46\% | -75.04\% |
| 0.2 | 350 | 67 | 283 |  | 0.156 | 0.024 | 0.132 | 1897 | 55.11\% | -51.73\% |
| 0.3 | 510 | 97 | 413 |  | 0.23 | 0.036 | 0.198 | 1758 | 43.75\% | -29.68\% |
| 0.4 | 660 | 125 | 535 |  | 0.31 | 0.048 | 0.26 | 1629 | 33.20\% | -9.02\% |
| 0.5 | 801 | 151 | 650 |  | 0.39 | 0.06 | 0.33 | 1509 | 23.39\% | 10.40\% |
| 0.6 | 934 | 175 | 758 |  | 0.47 | 0.072 | 0.4 | 1398 | 14.31\% | 28.71\% |
| 0.7 | 1059 | 198 | 861 |  | 0.55 | 0.083 | 0.46 | 1296 | 5.97\% | 46.05\% |
| 0.8 | 1177 | 219 | 958 |  | 0.62 | 0.095 | 0.53 | 1201 | -1.80\% | 62.28\% |
| 0.9 | 1288 | 238 | 1049 |  | 0.7 | 0.107 | 0.59 | 1113 | -8.99\% | 77.53\% |
| 1 | 1393 | 257 | 1136 |  | 0.78 | 0.119 | 0.66 | 1031 | -15.70\% | 92.09\% |
| 1.1 | 1492 | 273 | 1218 |  | 0.86 | 0.131 | 0.73 | 956 | -21.83\% | 105.83\% |
| 1.2 | 1586 | 289 | 1296 |  | 0.93 | 0.143 | 0.79 | 886 | -27.56\% | 118.72\% |
| 1.3 | 1674 | 304 | 1370 |  | 1.01 | 0.155 | 0.86 | 821 | -32.87\% | 130.93\% |
| 1.4 | 1758 | 318 | 1441 |  | 1.09 | 0.167 | 0.92 | 761 | -37.78\% | 142.58\% |
| 1.5 | 1838 | 330 | 1508 |  | 1.17 | 0.179 | 0.99 | 705 | -42.36\% | 153.54\% |
| 1.6 | 1914 | 342 | 1571 |  | 1.25 | 0.191 | 1.06 | 654 | -46.53\% | 163.94\% |
| 1.7 | 1985 | 354 | 1632 |  | 1.32 | 0.2 | 1.12 | 606 | -50.45\% | 173.79\% |
| 1.8 | 2053 | 364 | 1689 |  | 1.4 | 0.21 | 1.19 | 562 | -54.05\% | 183.22\% |
| 1.9 | 2118 | 374 | 1744 |  | 1.48 | 0.23 | 1.25 | 521 | -57.40\% | 192.23\% |
| 2 | 2180 | 383 | 1797 |  | 1.56 | 0.24 | 1.32 | 483 | -60.51\% | 200.69\% |
| 2.1 | 2238 | 392 | 1847 |  | 1.64 | 0.25 | 1.39 | 448 | -63.37\% | 208.74\% |
| 2.2 | 2294 | 400 | 1895 |  | 1.71 | 0.26 | 1.45 | 415 | -66.07\% | 216.51\% |
| 2.3 | 2348 | 407 | 1941 |  | 1.79 | 0.27 | 1.52 | 385 | -68.52\% | 223.86\% |
| 2.4 | 2399 | 414 | 1985 |  | 1.87 | 0.29 | 1.58 | 357 | -70.81\% | 230.93\% |
| 2.5 | 2447 | 421 | 2027 |  | 1.95 | 0.3 | 1.65 | 331 | -72.94\% | 237.73\% |

Table 37.14. Whiting 7.a. Management options Advice table. Weights in tonnes.

| Catch23 | Land23 | Dis23 | Basis | FCatch23 | FLand23 | FDis23 | SSB24 | dSSB | dTac |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 481 | 92 | 389 | FMSY | 0.22 | 0.034 | 0.185 | 1783 | 45.79\% | -33.70\% |
| 39 | 8 | 32 | FMSY x SSB(2023)/MSY Btrigger | 0.0164 | 0.0025 | 0.0139 | 2175 | 77.84\% | -94.59\% |
| 0 | 0 | 0 | $\mathrm{F}=0$ | 0 | 0 | 0 | 2211 | 80.79\% | -100\% |
| 1393 | 257 | 1136 | $\mathrm{F}=\mathrm{Fsq}$ | 0.78 | 0.119 | 0.66 | 1031 | -15.70\% | 92.09\% |
| 767 | 145 | 622 | $\mathrm{F}=\mathrm{Flim}$ | 0.37 | 0.057 | 0.31 | 1538 | 25.76\% | 5.69\% |
| 481 | 92 | 389 | $\mathrm{F}=\mathrm{Fpa}$ | 0.22 | 0.034 | 0.185 | 1783 | 45.79\% | -34\% |
| 355 | 68 | 287 | Min FMSY | 0.158 | 0.024 | 0.134 | 1893 | 54.78\% | -51.04\% |
| 28 | 5 | 23 | Min FMSY x SSB(2023)/MSY Btrigger | 0.0119 | 0.00181 | 0.01 | 2185 | 78.66\% | -96.12\% |
| 481 | 92 | 389 | Max FMSY | 0.22 | 0.034 | 0.185 | 1783 | 45.79\% | -33.70\% |
| 39 | 8 | 32 | Max FMSY x SSB(2023)/MSY Btrigger | 0.0164 | 0.0025 | 0.0139 | 2175 | 77.84\% | -94.59\% |
| 1149 | 214 | 935 | Stable SSB | 0.6 | 0.092 | 0.51 | 1223 | 0\% | 58.39\% |
| 850 | 160 | 690 | SSB * 1.2 | 0.42 | 0.064 | 0.35 | 1468 | 20.03\% | 17.20\% |
| 616 | 117 | 499 | -15\% TAC | 0.29 | 0.044 | 0.24 | 1666 | 36.22\% | -14.98\% |
| 721 | 136 | 585 | Stable TAC | 0.35 | 0.053 | 0.29 | 1573 | 28.62\% | 0\% |
| 834 | 157 | 677 | + 15\% TAC | 0.41 | 0.062 | 0.35 | 1481 | 21.10\% | 14.98\% |

### 37.14 Figures



Figure 37.1. Whiting 7.a. Working group estimates of International landings and discards.


Figure 37.2. Whiting 7.a. Landings and discards by fleet.


Figure 37.3. Whiting 7.a. discard length-frequency by national fleets for the OTB_CRU métier. Note due to low levels of retained catch, and hence low sampling, these data are not presented. Not updated at WGCSE, 2021.


Figure 37.4. Whiting 7.a. Proportion of discards by age (left) and year (right).


Figure 37.5. Whiting 7.a. Smoothed Stock Weights (Three year running average).


Figure 37.6. Whiting 7.a. Log Standardized indices of tuning fleets by cohort.


Figure 37.7. Whiting 7.a. Selectivity-at-age in the Catch.


Figure 37.8. Whiting 7.a. Observed and Predicted index CPUE.


Figure 37.9. Whiting 7.a. Observed and Predicted catch.


Figure 37.10. Whiting 7.a. Retrospective analysis of the final ASAP run with Mohn's Rho calculation. Image shows >5 peels but calculation is based on five peels only.


Figure 37.11. Whiting 7.a. Stock Summary Plot. The thick black line represents the ASAP assessment. Standard deviations from ASAP are shaded grey. The thick black line in the catch plot represents the predicted catch from ASAP.

## Landings yield 2023

SSB 2024


Figure 37.12. Whiting 7.a. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.

# Whiting (Merlangius merlangus) in divisions 7.b-c and 7.e-k (southern Celtic Seas and eastern English Channel) 

## Type of assessment in 2022

This stock assessment was benchmarked in 2020 (ICES, 2020). The model has been changed to a stochastic State-Space Assessment Model (SAM) and detailed in the Stock Annex. An interbenchmark was also carried out for this stock in 2021 (ICES, 2021). The model input data were updated with additional discard data, re-estimated weights-at-age, and a revision to the allocation of sampling across catch. Reference points were revised accordingly.

## ICES advice applicable to 2022

ICES advises that when the MSY approach is applied, catches in 2022 should be no more than 4452 tonnes.

ICES advice applicable to 2021
ICES advises that when the EU multiannual plan (MAP) for Western Waters and adjacent waters is applied, catches in 2021 that correspond to the F ranges in the MAP are between 4458 tonnes and 5261 tonnes.

### 38.1 General

## Stock description and management units

The TAC for whiting is set for divisions 7.b, 7.c, 7.d, 7.e, 7.f, 7.g, 7.h, 7.j and 7.k. The assessment area does not correspond to the TAC area. Since the 2014 Benchmark (WKCELT), Whiting in 7.b,c are now assessed as part of 7.bc, e-k, while whiting in 7.d remain part of the WGNSSK assessment of the North Sea stock. Any management measures implemented for this stock should be consistent with the assessment area.


Red Boxes-TAC/Management Areas Blue Shading-Assessment Area.

The TAC for whiting 7.bc, e-k decreased from 19 184t (2019) to 8352t (2022). ICES official landings for whiting $7 . \mathrm{bc}$, $\mathrm{e}-\mathrm{k}$ in 2022 are 6153 t and estimated catch of 7197 t . Thus, the current TAC for whiting catches in the $7 . \mathrm{b}-\mathrm{ce}-\mathrm{k}$ stock area is not restrictive in the $7 . \mathrm{bc}, \mathrm{e}-\mathrm{k}$ assessment area.

TAC in 2022/109


## Landings obligation

Since 2017 the landings obligation (LO) has applied to this stock in accordance with Delegated Regulation (EC, 2016) superseded by (EU) 2019/22391. This implies that all catches of whiting in the Celtic Sea and Western Channel by those vessels must be landed. However, a $6 \%$ de minimus applies to bottom trawls using a mesh size of $\geq 80 \mathrm{~mm}$, as well as pelagic trawls and beam trawls using $80-119 \mathrm{~mm}$ mesh. There are also three specific technical measures in operation for vessels using bottom trawls or seines in the Celtic Sea Protection Zone.

A significant proportion of unwanted catch is above the Minimum Conservation Reference Size $(M C R S=27 \mathrm{~cm})$ in whiting, although discards are assessed by ICES to have reduced in 2019 to $14 \%$ from $48-17 \%$ for $2016-2018$ respectively. Whiting is also the least limiting stock for most fleets in a mixed-fishery context for the Celtic Sea, where cod is most commonly considered the choke species. In this context it is difficult to accurately predict the impact of the LO on Celtic Sea whiting.

### 38.2 The fishery in 2021

ICES officially reported landings for divisions $7 . b-c e-k$ and landings as used by the Working Group are given in Tables

Table 1 1. In addition, landings for 7.d are included for comparison to the 7.b-k TAC (i.e. management area).

The 7.bc, e-k whiting stock is primarily targeted by otter trawlers and to a lesser extent Scottish seines and beam trawls. An overview of landings by fleet is given in Table 2 and more generally effort trends in fleets catching whiting in the Celtic Sea is provided by STECF (STECF, 2018).

The spatial distribution of international otter trawl effort by country 2014-2018 is given in Figure 1. Irish OTB effort is primarily from within 7.g (the Smalls fishing grounds gadoid fishery) and

[^23]to a lesser extent 7.j and the Porcupine Bank (Nephrops fishery). In previous years, French landings have exhibited similar spatial and temporal focus around the Smalls.

### 38.3 Data

## Catch

A general data handling approach was agreed during the WKCELTIC 2020 preparatory data workshop and reviewed again during IBPCSWhiting (ICES, 2021) following an error in the data and codesed to allocate sampling to un-sampled catch data. Data is submitted to Intercatch (IC) by France, Ireland, Belgium, UK, Spain and the Netherlands. A standardized approach to international catch data exploration and QC is taken across the cod, haddock and whiting stocks in the form of a shared R markdown document ${ }^{2}$. In so far as is possible, the allocation of sampling to un-sampled métiers is likewise standardized across stocks using the same R markdown template and editing only where necessary.
Fishery-dependent data are therefore collated in InterCatch, but raised and documented outside of InterCatch using these shared open source R Markdown documents.
Raising of un-sampled catches to International CNAA is implemented using a simple hierarchy for available samples where priority was given to the same:
i. Country \& Season \& Year
ii. Season \& Year
iii. Year

With gears into to either: GNS_DEF, OTB_CRU, OTB_DEF, TBB_DEF and MIS_MIS.
Discard raising is likewise implemented where samples were missing by estimating ratios at three levels:
i. Year, country and gear
ii. Year and gear
iii. Year

The international catch numbers-at-age are given in Table 3 and Figure 2. It is possible to track the strong 1999 and 2013 year classes, but the strong 2009 recruitment is only apparent at some older ages. Generally, the proportion of un-sampled catch that requires raising is minimal Figure 3, this inevitably increased somewhat during Covid-19, but was still deemed moderate compared to historic data for many stocks. The age distribution has remained similar over time with the exception of periods where strong year classes pass through older ages. Discarding of age 2 and above highlights significant fishing mortality above the minimum conservation reference size (MCRS - 27 cm ).

While poorly represented in the survey data, the 0-group age class is incorporated into the assessment data to allow inclusion of 0-group indices, although catches at this age are not minimal in most years. Catch weights-at-age and mean weights-at-age in the catch are given in Table 4 and 5 respectively. Rivard corrected stock weights (Table 6) were derived as per methodology described in the stock annex. The stock weights are shown in Figure 4. There is some variability of stock weights particularly at older ages, but $0-5 \mathrm{yr}$ old corrected weights are relatively stable. There is a notable increase in the 0 -group stock weight from the previous year and drop in 1-

[^24]group so they appear to overlap. This not so apparent in the mean weights-at-age, but a downweighting of the 1 -group fish in the Rivard correction process due to a low value in 2020. In contrast 0 -group fish have returned to a more average value.

## Discards

The time-series of discard data was revised by WKCELTIC and are included in the assessment. Procedures for raising discards to international catch-at-age are summarised above and detailed in the stock annex. More accurate national data, and for more Member States, are now available through InterCatch. Historically, Irish and French OTB discards were simply raised to international landings to produce an estimate of discards-at-age.

A summary of discarding rates-at-age for the time-series 2003-2021 available in InterCatch is presented in Figure 5. The two main fleets exploiting whiting, FRA_OTB and IRL_OTB, have shown some downward trend in discarding in recent years. The remaining lesser métiers have remained largely constant over time. Numbers and weights by age and country for the most recent data year (2021) are given in Table 7 for both landings and discards.

Figure 6 presents the proportion of landings and discards. The data indicate that the proportion of young being landed, versus discarded, has increased significantly in the last four years suggesting a distinct shift towards landing more fish around the MCRS.

## Biological

Mean stock and catch weights-at-age data were calculated following the methodology described in the stock annex. Natural mortality is based on Lorenzen's model and thus a power function of catch weights-at-age.

Mortality Ogive

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Proportion Mature | 1.22 | 0.86 | 0.65 | 0.50 | 0.43 | 0.40 | 0.38 | 0.36 |

Maturity was historically knife-edge at age 2, but has been replaced at the Benchmark to a revised maturity ogive based on survey data.

Maturity Ogive

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Proportion Mature | 0 | 0.61 | 0.94 | 0.97 | 0.97 | 1 | 1 | 1 |

## Surveys

Two IBTS Q4 surveys, FR-EVHOE and IE-IGFS, have been combined to provide the survey index for the assessment of Celtic Sea whiting since the previous benchmark in 2014.

Issues with survey data gaps in particular, highlighted by WGCSE for review by the benchmark, led to significant work being undertaken to implement a modelling approach to survey index calculation. The approach selected was the VAST (Vector Autoregressive Spatio-Temporal) model (www.github.com/james-thorson/VAST).

Internal consistency is $>47 \%$ for all age classes above 0-group and ca. 70\% for ages 2:3 (Figure 7). However, following evaluation with different survey age ranges during assessment model fitting, the survey index was truncated to 0-2 (Table 8).

Log mean standardised indices are given in Figure 8. Plot by cohort the index has got quite noisy in recent years. Plotting by years shows a marked downward trend since 2015 across age classes.

## Commercial Ipue

An updated French commercial tuning fleet for whiting was made available (Table 9). The Working Document Laviale et al $2019^{3}$ details the issues raised by the old commercial tuning fleet and the work done to provide the updated French commercial tuning index. In summary, the list of species and the threshold used to select trips has been modified to better account for the fact that cod is no longer a target of these fisheries, but more a bytcatch of whiting and haddock directed fisheries. Moreover, the commercial tuning now accounts for both landings and discards. While the French otter trawl fleet generally accounts for ca. $45 \%$ of the landings it is ostensibly concentrated around the 7.d Channel area and eastern Celtic Sea.

### 38.4 Historical stock development

A State-space (SAM) assessment is used for this stock applying the settings as agreed at WKCELTIC. Runs are available at Stockassessment.org. The full time-series was used (19992021) with one survey index (VAST) and one commercial index (FRA-OTB-lpue). The settings are detailed within the stock annex.

## Data screening

The methodology agreed at WKCELTIC is implemented and documented as an R Markdown document and available on the WGCSE ${ }^{4}$ SharePoint site. For consistency, routine exploratory analysis was carried out in parallel as in previous years using FLR under R version 4.0.5. The packages FLCore 2.6.18, and FLXSA 2.6.4 and FLEDA 2.5 were used.

## Final update assessment

The final assessment was run as per https://www.stockassessment.org/set-Stock.php?stock=whg.7b-ce-k WGCSE22 RevRec.

Final model inputs and settings were:

[^25]| $\bullet$ | Full time-series of catch data(1999 to 2021, ages 0 to 7+) |
| :--- | :--- |
| • | Model-filled discards for ages 5-7+ in 1999-2002 |
| $\bullet$ | VAST Model index for ages 0-2 from IGFS:EVHOE 2003-2020 |
| $\bullet$ | French Commercial biomass index in Kg/Hr for 2000-2020 |
| $\bullet$ | Fishing mortality states were bound for ages 6+ |
| • | Catchability for ages 1+ were bound for the survey index |
| • | Default settings for remaining configuration |
|  | Observation error on the first age in the survey was estimated separately from the older ages |
|  | (i.e. ages 1-2 were bound). |

Fishing mortality-at-age and stock numbers-at-age are presented in Tables 10 and 11 respectively. Summary plots for SSB, Fbar and Recruitment are given in Figure 9. The last small pulse in recruitment in 2013 resulted in a small rise in SSB as harvestable 2-year olds in 2015. With poor recruitment since there has been little to bolster the stock or fishery since 2016.

Model fits to the data are presented in Figure 10. The overall fit to catch data are reasonably good and the IBTS survey observations are also oscillating around the SAM model. Fit for the commercial biomass index is less predictable however, possibly due in part to spatial coverage. The same patterns are reflected in the residuals presented in Figure 11 with $0-2 y r$ old indices slightly higher than predicted for last year and catch observations conversely appearing slightly lower for 0-group and ages $4+$.

## Comparison with previous assessments

Preliminary runs showed improving precision over the previous year's assessment, but with a continuing retrospective bias highlighted by the Mohn's Rho values (Figure 11). Spawning-stock biomass being somewhat over estimated annually and fishing mortality being consistently revised upwards. The positive residual patterns for the FRA-OTB biomass index was discussed and exploratory runs presented excluding this index, but felt a significant deviation to implement outside a benchmark process.

## State of the stock

Trends in landings, $\mathrm{F}_{(2-5)}, \mathrm{SSB}$, and recruitment are presented in Table 12. For the recent timeseries, SSB displays a peak biomass in 2010 following relatively strong recruitment from the 2008-2009 year classes. Again in 2014-2015 following the 2013 recruitment.

Fishing mortality ( $\mathrm{Fbar}^{\text {}}$ ) increased between 2012 and 2016, but is now assessed to be just below Flim but above FmSy. SSB is estimated to be increasing slowly, but still below agreed reference points.

### 38.5 Short-term projections

The short-term projections were carried out in SAM (stockassessment.org) as described in the stock annex. However, given the apparent bias a revision to the recruitment assumption was used span only the recent seven years from 2015-2021 rather than starting in 2010. This would remove the moderately high 2013 recruitment point producing a more conservative estimate of recruitment and therefore SSB in the forecast.

Whiting in the Celtic Sea, as with many gadoid fisheries, is heavily reliant on younger age classes and therefore recruitment. Recruitment is highly sporadic and thus the span over which a mean or median recruitment assumption is taken for the intermediate year is important. Historically this was taken as GM for the time-series minus the last year. Following discussion at WKCELTIC, and further the ADG, this was revised to median since 2010 which covers the more modern history of the stock. The median resampled recruitment then from 2010-2021 in 000s was estimated as 507319 and with the final 2015-2021 year range as 400108 . This was used as 0 -group numbers in the forecast for 2022-2023.

Table 13 gives the management options table. Given the probability of SSB being below Blim of $>50 \%$ for all scenarios other than zero catch, $\mathrm{F}=0$ is advised. Fishing at 0 gives a $40 \%$ likelihood of being below Blim with a spawning-stock biomass forecast of $38,109 \mathrm{t}$. The assumed recruitment in 2022 and 2023 used in the forecast constitutes a significant part ( $49.1 \%$ ) of the projected SSB in 2024 (Figure 13).

The basis for the catch forecast is given in Table 14. Whiting is aligned with the other benchmarked species (cod and haddock) in this mixed fishery. A catch constraint is generated by taking the whiting catch predicted by Mixed Fish ( 9240 t ) for F at Haddock Fmš. The resulting Fmix ( 0.571 ) was then used as the F assumption in the intermediate year (2022).

### 38.6 MSY evaluations and Biological reference points

ICES carried out an evaluation of MSY and PA reference points for this stock at IBPCSWhiting (ICES, 2021). The results are summarised below:

## Reference points

| Reference <br> Point | IBPCSWhiting 2021 <br> Value | WKCELTIC 2020 <br> Value | Rationale |
| :---: | :---: | :---: | :---: |
| MSY Btrigger | 50818 t | 47963 t | $B_{p a}$ |
| Fmsy | 0.375 | 0.4 | From EqSim with segmented regression and fixed breakpoint ( $\mathrm{B}_{\mathrm{lim}}$ ) capped to $\mathrm{F}_{\mathrm{p} 0.5}$. |
| FmsyLower | 0.315 | 0.332 | Median lower point estimates of ( $\mathrm{F}_{05}$ ) |
| FmsyUpper | 0.375 | 0.4 | $\mathrm{F}_{\mathrm{p} .05}$ |
| Blim | 36571 t | 34516 t | Bloss; lowest observed SSB (2008) from which stock recovery was observed. |
| Bpa | 50818 t | 47963 t | Blim combined with the assessment error; $\mathrm{Blim}_{\times \exp }$ (1.645 $\times \sigma$ ); $\sigma=0.20$ (default setting) |
| Flim | 0.64 | 0.89 | F with 50\% probability of SSB less than Blim |

### 38.7 Management plans

The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to demersal stocks including whiting in ICES divisions 7.b-ce-k.

### 38.8 Uncertainties and bias in assessment and forecast

## Sampling

Sampling levels of the landed catch for recent years are considered to be sufficient to support current assessment approaches. Revised time-series data for 2002-2021 are now included in the assessment with sampling from more countries which should give greater accuracy. While the overall SOP checks have invariably been $\leq 3 \%$, any difference in the sampled catch-at-age going into the assessment vs those coming out will cause concern. Rather than correct the national data provided, a SOP correction is applied as part of the revised raising procedures outlined above and the stock annex.

## Ageing

Cohort tracking in the landings-at-age matrix appears consistent up to age 6. Tracking deteriorates at older ages.

## Discards

Discarding has been major feature of most fleets catching whiting in the Celtic Sea. Sampling coverage of discarding has improved over time particularly since 2004. Discard estimates for the UK and Belgium are now included along with those of Ireland and France.

## Selectivity

Square-mesh panels were introduced in the second half of 2012 to reduce catches and discards of smaller whiting and haddock. The current assessment does not show an obvious reduction in F-at-age since the introduction of this TCM.

## Surveys

The survey indices for whiting are prone to some year effects. However, cohort tracking for the $1+$ fish is consistent and has improved further using the VAST modelling approach. There is a noticeable downward trend since 2016 in the indices plotted by year and higher noise when plotted by cohort (Figure 8).

## Misreporting

The level of misreporting for this stock is not known. Underreporting has previously been considered unlikely to be a significant source of unaccounted mortality in the assessment because the TAC has been in excess of recent catches.

### 38.9 Recommendation for next benchmark

The survey indices were truncated from 0-5 year olds down to $0-2$ year olds as part of model fit optimization. This should be revisited again to ensure the model is not over fitting to the catch data. Commercial tuning was only available from France at the recent benchmark, whose fleet have a different spatial extent to that of Ireland, the other main country involved in the fishery. Potential to extend the coverage of the commercial tuning index should be examined otherwise whether the SAM fit to the LPUE index is reasonable.

### 38.10 Management considerations

Catches and SSB in 7.b, c, e-k whiting fluctuate considerably depending on year-class strength. The 2008 and 2009 year classes were above average with 2013 being third highest in the timeseries. These contributed to catches and SSB in the short term but the upturn in catches and SSB was short lived as recruitment is episodic and SSB is now below all reference points.

Discarding in this stock for different fleets is substantial and highly variable depending on gear and year-class strength. High levels of discarding for a species like whiting reduce the longer term yields one might expect so efforts to improve selection and reduce discards in the mixed fishery should be encouraged. ICES notes the introduction of square-mesh panels in all trawl fisheries operating in ICES divisions 7.fg. It is important that these measures are fully implemented and their effectiveness in reducing discards and the impact on commercial catches is monitored and evaluated. Further gear modifications to increase the likelihood of small whiting passing through the gear, such as introduction of larger minimum mesh sizes, separator panels, or grids may be needed.

Ireland has the only directed fishery for whiting which is part of mixed fishery throughout the Celtic Sea, as well as bycatch within Nephrops fisheries. Discard rates are high as a consequence of the low market value of the species, particularly at smaller sizes. High-grading above the MCRS to some extent is also prevalent in most fisheries.

From the 1 February to the 31 March fishing activity has been prohibited within ICES rectangles: 30E4, 31E4, 32E3 (excluding within six nautical miles from the baseline) annually since 2005 to protect the cod stock.

There have been major changes in fleet dynamics over the period of the assessment. Effort in the French gadoid fleet has been declining since 1999, but the effort has fluctuated in recent years due to the way the effort series is derived. Irish otter-trawl effort in $7 . \mathrm{b}-\mathrm{k}$ has also declined slightly over the time-series.

The full impact of the Landings Obligation is complex and unknown as yet and will depend on whether there is a measurable impact on discarding behaviour or whether variable practices continue and simply data becomes more reliable (for a summary of issues see http://www.discardless.eu/media/results/Celtic Sea Year2.pdf).

### 38.11 References

EC. 2015. Commission Delegated Regulation (EU) 2015/2438 of 12 October 2015 establishing a discard plan for certain demersal fisheries in north-western waters.

EC. 2016. Commission Delegated Regulation (EU) 2016/2375 of 12 October 2016 establishing a discard plan for certain demersal fisheries in north-western waters.

ICES. 2016a. Report of the Workshop to consider Fmsy ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

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ICES. 2021. Inter-Benchmark Protocol on Celtic Seas Whiting (IBPCSWhiting 2021). ICES Scientific Reports. 3:103. https://doi.org/10.17895/ices.pub. 8718.

### 38.12 Tables

Table 1. Whiting in Divisions 7.bc,e-k. Nominal Landings (t) as reported to ICES, and total landings as used by the Working Group.

| Year | Official ICES Landings |  |  |  |  |  |  | Used by WG |  |  | 7.bc,e-k Catch + |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | BEL | FRA | IRL | UK_EW | Others | Total | Unallocated | WG Total | Dicards | Catch | 7.d Landings | TAC |
| 1998 | 479 | 11748 | 5549 | 1755 | 179 | 19710 | - | - | - | - |  |  |
| 1999a | 448 | 16418 | 6013 | 1354 | 27 | 24260 | 4082 | 20178 | 5420 | 25598 | 31401 |  |
| 2000 | 194 | 9186 | 5358 | 1255 | 39 | 16032 | 387 | 15645 | 4400 | 20045 | 26117 |  |
| 2001 | 171 | 7317 | 5365 | 948 | 31 | 13832 | 640 | 13192 | 9877 | 23070 | 29684 |  |
| 2002 | 149 | 7548 | 5718 | 847 | 35 | 14297 | 657 | 13640 | 7336 | 20977 | 26338 |  |
| 2003 | 129 | 5989 | 4516 | 763 | 21 | 11418 | 321 | 11097 | 3559 | 14656 | 21661 |  |
| 2004 | 180 | 4874 | 4350 | 587 | 132 | 10123 | -66 | 10189 | 6481 | 16670 | 21953 |  |
| 2005 | 218 | 5913 | 5774 | 482 | 136 | 12523 | 312 | 12211 | 6700 | 18911 | 23812 |  |
| 2006 | 128 | 4710 | 4570 | 413 | 129 | 9951 | 291 | 9660 | 12031 | 21691 | 25440 |  |
| 2007 | 127 | 3574 | 4864 | 576 | 86 | 9226 | 139 | 9087 | 8456 | 17543 | 20934 | 19900 |
| 2008 | 121 | 3072 | 2406 | 620 | 35 | 6255 | 395 | 5860 | 2880 | 8740 | 11933 | 19900 |
| 2009 | 87 | 2814 | 2798 | 827 | 25 | 6551 | 38 | 6513 | 4101 | 10614 | 17183 | 16950 |
| 2010 | 102 | 3463 | 4330 | 798 | 85 | 8779 | 191 | 8588 | 3008 | 11596 | 17729 | 14407 |
| 2011 | 100 | 4312 | 4752 | 740 | 174 | 10077 | 593 | 9484 | 1954 | 11438 | 16902 | 16658 |
| 2012 | 170 | 3710 | 5841 | 764 | 141 | 10627 | 439 | 10188 | 2449 | 12637 | 16234 | 19053 |


| Year | Official ICES Landings |  |  |  |  |  |  | Used by WG |  |  | 7.bc,e-k Catch + |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | BEL | FRA | IRL | UK_EW | Others | Total | Unallocated | WG Total | Dicards | Catch | 7.d Landings | TAC |
| 2013 | 226 | 4006 | 6888 | 907 | 92 | 12119 | 188 | 11931 | 2512 | 14443 | 18700 | 24500 |
| 2014 | 222 | 4928 | 6874 | 1062 | 35 | 13121 | 274 | 12847 | 3977 | 16824 | 19954 | 19162 |
| 2015 | 152 | 5634 | 6437 | 828 | 97 | 13149 | -25 | 13174 | 6101 | 19275 | 19954 | 17742 |
| 2016 | 186 | 6294 | 7700 | 892 | 39 | 15110 | -69 | 15179 | 7278 | 22457 | 26187 | 22778 |
| 2017 | 102 | 5256 | 6296 | 607 | 32 | 12293 | 600 | 11693 | 4505 | 17098 | 17780 | 27500 |
| 2018 | 103 | 3666 | 4628 | 592 | 31 | 9019 | 246 | 8773 | 1495 | 10268 | 12625 | 22213 |
| 2019 | 73 | 3203 | 2599 | 487 | 126 | 6488 | 946 | 5542 | 752 | 6294 | 9393 | 19184 |
| 2020* | 82 | 2669 | 2650 | 336 | 90 | 5827 | 104 | 5931 | 1266 | 7197 | 1970 | 10863 |
| 2021* | 82 | 2666 | 2915 | 352 | 66 | 6082 | 71 | 6153 | 1224 | 7377 | 2454 | 12259 |

"Provisional data.
${ }^{\text {a }}$ French Official landings not available, not updated.

Table 2. Whiting in Divisions 7.b-ce-k. Landings (t) by fleet.

| LANDINGS | Others | ОTB | SSC | TBB | Total | \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BEL | 0 | 14 | 1 | 65 | 81 | 1\% |
| FRA | 120 | 2561 | 0 | 0 | 2681 | 44\% |
| IRL | 36 | 1884 | 1014 | 38 | 2973 | 48\% |
| UK | 39 | 240 | 10 | 65 | 354 | 6\% |
| Others | 1 | 1 | 63 | 0 | 65 | 1\% |
| Total | 196 | 4700 | 1088 | 168 | 6153 | 100\% |
|  | 3\% | 76\% | 18\% | 3\% | 100\% |  |

Table 3. Whiting in divisions 7.bc,e-k. The strong 1999 year class is distinct in both the catch and landings data, with evidence of the strong 2009 and 2013 year classes appearing at older ages. Catch numbers-at-age ('000).

| 1999 | 2021 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 7 |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| 5370.0 | 20744.1 | 25957.7 | 14662.4 | 8744.8 | 8987.8 | 6670.2 | 1498.7 |
| 8176.3 | 26561.7 | 26303.7 | 12529.9 | 6122.5 | 2605.9 | 2100.9 | 2424.3 |
| 8795.0 | 26105.8 | 51390.6 | 13715.2 | 5317.1 | 2049.0 | 763.1 | 627.3 |
| 4568.6 | 13387.4 | 34319.6 | 24356.6 | 5968.2 | 1057.6 | 291.6 | 111.0 |
| 13563.8 | 20962.0 | 34625.2 | 14881.0 | 15187.9 | 2698.4 | 369.8 | 4.2 |
| 35663.6 | 20301.8 | 60277.3 | 30276.2 | 15671.1 | 6833.3 | 541.1 | 77.0 |
| 5540.3 | 33978.7 | 44751.0 | 18055.2 | 8245.2 | 6434.5 | 2651.8 | 126.4 |
| 13472.7 | 16455.6 | 8974.9 | 9465.6 | 4559.3 | 2821.7 | 4419.0 | 634.9 |
| 926.1 | 10977.9 | 29863.4 | 22446.5 | 6347.2 | 2601.3 | 821.3 | 1016.0 |
| 1430.2 | 10540.5 | 14640.9 | 10936.2 | 3775.9 | 865.0 | 220.4 | 89.7 |
| 809.6 | 6124.2 | 17584.6 | 10350.5 | 3958.6 | 1266.2 | 248.0 | 78.2 |
| 495.8 | 12773.2 | 15669.6 | 14991.2 | 4803.2 | 1207.5 | 283.2 | 104.8 |
| 559.8 | 4153.3 | 15044.6 | 12540.0 | 6502.9 | 1626.1 | 375.8 | 102.2 |
| 3798.5 | 6573.8 | 9025.6 | 15864.1 | 7519.9 | 2653.6 | 605.5 | 134.1 |
| 770.0 | 3346.0 | 8808.5 | 7320.9 | 12392.4 | 4809.3 | 1054.1 | 294.5 |
| 133.9 | 14770.7 | 6808.8 | 7768.4 | 6684.4 | 7574.4 | 1746.7 | 301.5 |
| 4647.2 | 5651.8 | 32558.2 | 7710.9 | 6203.1 | 2815.4 | 3111.9 | 650.8 |
| 2074.9 | 10980.8 | 13651.0 | 33791.0 | 5935.6 | 3085.0 | 1079.9 | 1193.1 |
| 933.6 | 2840.6 | 12286.9 | 7615.0 | 11764.8 | 2010.4 | 771.2 | 282.8 |
| 1803.3 | 2888.6 | 8804.0 | 7711.5 | 3749.0 | 3979.7 | 575.2 | 219.5 |
| 93.0 | 3025.8 | 4713.2 | 4371.8 | 3044.0 | 1017.7 | 745.5 | 75.3 |
| 1417.2 | 3684.5 | 8679.1 | 3972.3 | 1534.9 | 758.1 | 219.9 | 126.2 |
| 48.7 | 4556.3 | 5814.8 | 8063.6 | 1688.6 | 422.1 | 124.2 | 34.9 |

Table 4. Whiting in divisions 7.bc,e-k. Catch weights-at-age (Tons).

| 1999 | 2021 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 7 |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |
| 603.1 | 2588.8 | 6681.7 | 4496.2 | 6085.5 | 1416.0 | 250.9 | 2.5 |
| 748.8 | 3135.3 | 10982.9 | 7433.9 | 4674.5 | 2629.2 | 204.7 | 32.8 |
| 229.2 | 3989.7 | 8773.7 | 5791.2 | 3439.3 | 2739.9 | 1143.7 | 66.1 |
| 467.9 | 2433.0 | 2529.3 | 3491.9 | 2416.2 | 1601.7 | 1673.4 | 392.8 |
| 42.6 | 1403.4 | 5695.9 | 6364.0 | 2407.3 | 1230.4 | 374.8 | 345.0 |
| 54.0 | 1298.9 | 3080.3 | 3088.6 | 1658.0 | 424.3 | 159.9 | 76.6 |
| 54.7 | 844.6 | 3662.2 | 3466.6 | 1780.3 | 838.7 | 147.6 | 60.2 |
| 20.4 | 1932.4 | 3935.0 | 5696.2 | 2404.7 | 684.4 | 183.8 | 59.5 |
| 17.7 | 716.1 | 3557.8 | 4520.9 | 3559.4 | 1104.6 | 263.8 | 94.0 |
| 217.0 | 677.7 | 2014.9 | 6407.2 | 4094.8 | 1945.2 | 462.3 | 130.5 |
| 35.2 | 493.6 | 1860.8 | 2657.4 | 6926.6 | 3052.4 | 905.4 | 225.0 |
| 6.5 | 2046.5 | 1742.6 | 3076.4 | 3667.0 | 5455.4 | 1365.1 | 295.4 |
| 258.3 | 682.6 | 7744.0 | 2961.8 | 3345.0 | 2059.2 | 2125.4 | 461.4 |
| 89.3 | 1355.3 | 2896.0 | 12098.1 | 3279.0 | 2093.7 | 777.7 | 871.1 |
| 39.6 | 409.9 | 2885.4 | 3015.1 | 6421.4 | 1450.4 | 689.8 | 256.3 |
| 98.4 | 364.0 | 1742.9 | 2878.6 | 2260.6 | 3064.2 | 516.5 | 221.3 |
| 5.8 | 496.3 | 1608.6 | 2056.9 | 1890.0 | 794.9 | 625.9 | 80.2 |
| 27.4 | 609.7 | 2686.0 | 2080.6 | 964.8 | 536.8 | 174.2 | 117.3 |
| 2.3 | 413.3 | 1672.8 | 3742.6 | 1092.5 | 327.2 | 88.5 | 37.0 |

Table 5. Whiting in divisions 7.bc,e-k. Mean catch weights-at-age (kg).

|  | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1999 | 0.0271 | 0.1331 | 0.2216 | 0.3412 | 0.4274 | 0.4402 | 0.4963 | 0.623 |
| 2000 | 0.0314 | 0.069 | 0.2204 | 0.3955 | 0.5053 | 0.563 | 0.5804 | 0.5868 |
| 2001 | 0.0315 | 0.1116 | 0.1853 | 0.3778 | 0.5293 | 0.6335 | 0.76 | 0.7775 |
| 2002 | 0.0272 | 0.0965 | 0.1966 | 0.3506 | 0.5315 | 0.7069 | 0.8249 | 1.0133 |
| 2003 | 0.0445 | 0.1235 | 0.1930 | 0.3021 | 0.4007 | 0.5248 | 0.6786 | 0.6038 |
| 2004 | 0.0210 | 0.1544 | 0.1822 | 0.2455 | 0.2983 | 0.3848 | 0.3783 | 0.4263 |
| 2005 | 0.0414 | 0.1174 | 0.1961 | 0.3207 | 0.4171 | 0.4258 | 0.4313 | 0.5232 |
| 2006 | 0.0347 | 0.1479 | 0.2818 | 0.3689 | 0.5300 | 0.5676 | 0.3787 | 0.6186 |
| 2007 | 0.0460 | 0.1278 | 0.1907 | 0.2835 | 0.3793 | 0.4730 | 0.4563 | 0.3395 |
| 2008 | 0.0377 | 0.1232 | 0.2104 | 0.2824 | 0.4391 | 0.4905 | 0.7256 | 0.8543 |
| 2009 | 0.0675 | 0.1379 | 0.2083 | 0.3349 | 0.4497 | 0.6624 | 0.5952 | 0.7689 |
| 2010 | 0.0411 | 0.1513 | 0.2511 | 0.3800 | 0.5007 | 0.5668 | 0.6489 | 0.5674 |
| 2011 | 0.0316 | 0.1724 | 0.2365 | 0.3605 | 0.5474 | 0.6793 | 0.7019 | 0.9197 |
| 2012 | 0.0571 | 0.1031 | 0.2232 | 0.4039 | 0.5445 | 0.7331 | 0.7635 | 0.9731 |
| 2013 | 0.0457 | 0.1475 | 0.2113 | 0.3630 | 0.5589 | 0.6347 | 0.8589 | 0.7641 |
| 2014 | 0.0484 | 0.1386 | 0.2559 | 0.3960 | 0.5486 | 0.7202 | 0.7815 | 0.9798 |
| 2015 | 0.0556 | 0.1208 | 0.2379 | 0.3841 | 0.5392 | 0.7314 | 0.6830 | 0.7089 |
| 2016 | 0.0431 | 0.1234 | 0.2121 | 0.3580 | 0.5524 | 0.6787 | 0.7201 | 0.7301 |
| 2017 | 0.0424 | 0.1443 | 0.2348 | 0.3959 | 0.5458 | 0.7214 | 0.8945 | 0.9065 |
| 2018 | 0.0546 | 0.1260 | 0.1980 | 0.3733 | 0.6030 | 0.7700 | 0.8980 | 1.0080 |
| 2019 | 0.0625 | 0.1640 | 0.3413 | 0.4705 | 0.6209 | 0.7811 | 0.8396 | 1.0646 |
| 2020 | 0.0193 | 0.1655 | 0.3095 | 0.5238 | 0.6286 | 0.7081 | 0.7922 | 0.9297 |
| 2021 | 0.0473 | 0.0907 | 0.2877 | 0.4641 | 0.6470 | 0.7751 | 0.7123 | 1.0611 |

Table 6. Whiting in divisions 7.bc,e-k. Q1 Stock weights-at-age (kg) from Rivard corrected annual mean catch weights.

|  | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1999 | 0.01700 | 0.10340 | 0.16590 | 0.28040 | 0.37240 | 0.38340 | 0.46740 | 0.62300 |
| 2000 | 0.01670 | 0.04320 | 0.17130 | 0.29600 | 0.41520 | 0.49050 | 0.50550 | 0.58680 |
| 2001 | 0.01800 | 0.05920 | 0.11310 | 0.28860 | 0.45750 | 0.56580 | 0.65410 | 0.77750 |
| 2002 | 0.01280 | 0.05510 | 0.14810 | 0.25490 | 0.44810 | 0.61170 | 0.72290 | 1.01330 |
| 2003 | 0.02390 | 0.05800 | 0.13650 | 0.24370 | 0.37480 | 0.52810 | 0.69260 | 0.60380 |
| 2004 | 0.00890 | 0.08290 | 0.15000 | 0.21770 | 0.30020 | 0.39270 | 0.44560 | 0.42630 |
| 2005 | 0.02190 | 0.04970 | 0.17400 | 0.24170 | 0.32000 | 0.35640 | 0.40740 | 0.52320 |
| 2006 | 0.01810 | 0.07820 | 0.18190 | 0.26900 | 0.41230 | 0.48660 | 0.40160 | 0.61860 |
| 2007 | 0.02810 | 0.06660 | 0.16790 | 0.28260 | 0.37410 | 0.50070 | 0.50890 | 0.33950 |
| 2008 | 0.01970 | 0.07530 | 0.16400 | 0.23210 | 0.35280 | 0.43130 | 0.58580 | 0.85430 |
| 2009 | 0.04510 | 0.07210 | 0.16020 | 0.26540 | 0.35640 | 0.53930 | 0.54030 | 0.76890 |
| 2010 | 0.02010 | 0.10110 | 0.18610 | 0.28130 | 0.40950 | 0.50490 | 0.65560 | 0.56740 |
| 2011 | 0.01750 | 0.08420 | 0.18920 | 0.30090 | 0.45610 | 0.58320 | 0.63070 | 0.91970 |
| 2012 | 0.03550 | 0.05710 | 0.19620 | 0.30910 | 0.44300 | 0.63350 | 0.72020 | 0.97310 |
| 2013 | 0.02620 | 0.09180 | 0.14760 | 0.28460 | 0.47510 | 0.58790 | 0.79350 | 0.76410 |
| 2014 | 0.03060 | 0.07960 | 0.19430 | 0.28930 | 0.44630 | 0.63440 | 0.70430 | 0.97980 |
| 2015 | 0.03730 | 0.07650 | 0.18160 | 0.31350 | 0.46210 | 0.63340 | 0.70140 | 0.70890 |
| 2016 | 0.02360 | 0.08280 | 0.16010 | 0.29180 | 0.46060 | 0.60490 | 0.72570 | 0.73010 |
| 2017 | 0.02460 | 0.07890 | 0.17020 | 0.28980 | 0.44200 | 0.63130 | 0.77920 | 0.90650 |
| 2018 | 0.03150 | 0.07310 | 0.16900 | 0.29610 | 0.48860 | 0.64830 | 0.80490 | 1.00800 |
| 2019 | 0.03840 | 0.09460 | 0.20740 | 0.30520 | 0.48140 | 0.68630 | 0.80400 | 1.06460 |
| 2020 | 0.00890 | 0.10170 | 0.22530 | 0.42280 | 0.54380 | 0.66310 | 0.78660 | 0.92970 |
| 2021 | 0.05350 | 0.04180 | 0.21820 | 0.37900 | 0.58210 | 0.69800 | 0.71020 | 1.06110 |

Table 7. Whiting in divisions 7.e-k. Summary of landings and discard data for 2021 provided to the Working Group.

| weight in tonnes |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DIscards | COUNTRY | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ | Grand total |
|  | Belgium | 0.1 | 25.8 | 41.9 | 24.2 | 3.9 | 0.2 | 0.2 | 0.0 | 96.4 |
|  | France | 1.7 | 177.2 | 334.0 | 198.5 | 22.1 | 0.9 | 1.4 | 0.0 | 735.9 |
|  | Ireland | 0.0 | 83.5 | 112.1 | 65.9 | 14.9 | 0.6 | 1.0 | 0.0 | 277.9 |
|  | UK (England) | 0.4 | 9.8 | 33.1 | 13.8 | 1.4 | 0.4 | 0.1 | 0.0 | 59.1 |
|  | Other | 0.1 | 14.5 | 23.6 | 13.8 | 2.1 | 0.1 | 0.1 | 0.0 | 54.3 |
|  | Total | 2.3 | 310.7 | 544.8 | 316.3 | 44.4 | 2.3 | 2.8 | 0.0 | 1223.6 |
| Landings | Belgium | 0.0 | 1.2 | 14.5 | 45.3 | 13.7 | 4.2 | 1.1 | 0.5 | 80.6 |
|  | France | 0.0 | 92.4 | 725.2 | 1428.0 | 327.2 | 87.0 | 18.7 | 2.1 | 2680.6 |
|  | Ireland | 0.0 | 3.4 | 315.8 | 1742.6 | 623.6 | 200.8 | 56.0 | 30.6 | 2972.8 |
|  | UK (England) | 0.0 | 5.2 | 62.2 | 166.0 | 67.8 | 27.2 | 8.4 | 3.3 | 340.0 |
|  | Other | 0.0 | 0.4 | 10.3 | 44.4 | 15.8 | 5.7 | 1.5 | 0.5 | 78.6 |
|  | Total | 0.0 | 102.6 | 1128.0 | 3426.3 | 1048.1 | 324.9 | 85.7 | 37.0 | 6152.5 |

Number in 000's

| Discards | Country |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | Grand Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Belgium | 2.9 | 370.7 | 213.6 | 92.3 | 12.9 | 1.1 | 0.7 | 0.0 | 694.3 |
|  | France | 34.1 | 2124.7 | 1542.6 | 733.7 | 72.4 | 5.5 | 3.9 | 0.0 | 4517.0 |


| weight in tonnes |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ireland | 0.0 | 1377.3 | 622.7 | 257.8 | 50.1 | 4.4 | 2.8 | 0.0 | 2315.2 |
|  | UK (England) | 9.6 | 71.1 | 171.9 | 52.2 | 3.5 | 0.8 | 0.2 | 0.0 | 309.4 |
|  | Other | 2.0 | 204.1 | 118.5 | 52.2 | 6.8 | 0.6 | 0.4 | 0.0 | 384.6 |
|  | Total | 48.7 | 4147.9 | 2669.3 | 1188.2 | 145.7 | 12.5 | 8.0 | 0.1 | 8220.4 |
| Landings | Belgium | 0.0 | 4.7 | 40.4 | 91.3 | 20.2 | 5.4 | 1.6 | 0.5 | 163.9 |
|  | France | 0.0 | 368.7 | 2106.6 | 2851.3 | 451.8 | 90.4 | 21.9 | 4.1 | 5894.9 |
|  | Ireland | 0.0 | 11.2 | 765.9 | 3466.1 | 948.6 | 272.1 | 81.2 | 27.2 | 5572.3 |
|  | UK (England) | 0.0 | 22.1 | 201.5 | 377.6 | 100.2 | 35.4 | 9.6 | 2.6 | 749.0 |
|  | Other | 0.0 | 1.7 | 31.0 | 89.1 | 22.0 | 6.4 | 1.9 | 0.4 | 152.5 |
|  | Total | 0.0 | 408.3 | 3145.4 | 6875.4 | 1542.8 | 409.6 | 116.2 | 34.8 | 12532.7 |

Table 8. Whiting in divisions 7.bc,e-k. Combined (IE-IGFS and FR_EVHOE) VAST recruitment survey index for age groups 0-2 (No/Km²).

| IGFSEVHOE No/Hr |  |  |  |
| :---: | :---: | :---: | :---: |
| Age |  |  |  |
|  |  | 1 | 2 |
| 2003 | 46205.02 | 40655.983 | 12533.098 |
| 2004 | 157691.534 | 28393.568 | 7434.628 |
| 2005 | 37035.107 | 31625.849 | 5542.815 |
| 2006 | 73694.246 | 19836.628 | 7491.025 |
| 2007 | 231611.031 | 27207.135 | 4406.281 |
| 2008 | 161259.028 | 38695.505 | 9387.597 |
| 2009 | 231100.591 | 62799.181 | 10512.46 |
| 2010 | 21900.202 | 45146.326 | 19875.118 |
| 2011 | 63739.781 | 15863.561 | 24673.54 |
| 2012 | 39744.354 | 17373.611 | 9713.187 |
| 2013 | 236267.312 | 11165.44 | 7168.082 |
| 2014 | 29265.971 | 38110.41 | 6455.985 |
| 2015 | 72084.394 | 14574.46 | 16569.482 |
| 2016 | 73908.284 | 18872.027 | 10601.144 |
| 2017 | 71216.312 | 8958.637 | 4255.448 |
| 2018 | 84964.539 | 5825.47 | 2431.677 |
| 2019 | 52307.438 | 21968.034 | 3260.777 |
| 2020 | 41224.611 | 7403.799 | 4252.018 |
| 2021 | 47267.467 | 17694.33 | 6953.716 |

Table 9. Whiting in divisions 7.bc,e-k. FRA-OTB commercial biomass index ( $\mathrm{Kg} / \mathrm{Hr}$ ).

|  | $\mathrm{Kg} / \mathrm{Hr}$ |
| :---: | :---: |
| 2000 | 38.10363867 |
| 2001 | 20.72032437 |
| 2002 | 19.72791635 |
| 2003 | 15.04609422 |
| 2004 | 15.08119522 |
| 2005 | 24.65779777 |
| 2006 | 24.11897529 |
| 2007 | 14.66450994 |
| 2008 | 11.05968544 |
| 2009 | 11.14466828 |
| 2010 | 14.68285952 |
| 2011 | 13.01333083 |
| 2012 | 10.45746782 |
| 2013 | 13.16969924 |
| 2014 | 19.60473794 |
| 2015 | 20.3091624 |
| 2016 | 25.69082281 |
| 2017 | 25.06670645 |
| 2018 | 22.20498986 |
| 2019 | 22.43714973 |
| 2020 | 18.57563875 |
| 2021 | 16.47383047 |

Table 10. Whiting in divisions 7.b, c, e-k. Fishing mortality (F)-at-age. Fbar range is 2-5.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 0.009 | 0.096 | 0.425 | 0.656 | 0.944 | 1.189 | 1.573 | 1.573 |
| 2000 | 0.008 | 0.091 | 0.389 | 0.627 | 0.968 | 1.240 | 1.756 | 1.756 |
| 2001 | 0.008 | 0.088 | 0.361 | 0.581 | 1.047 | 1.555 | 2.354 | 2.354 |
| 2002 | 0.008 | 0.076 | 0.288 | 0.403 | 0.703 | 1.094 | 1.903 | 1.903 |
| 2003 | 0.009 | 0.090 | 0.330 | 0.384 | 0.601 | 0.978 | 1.547 | 1.547 |
| 2004 | 0.014 | 0.156 | 0.619 | 0.694 | 0.697 | 0.865 | 1.303 | 1.303 |
| 2005 | 0.012 | 0.149 | 0.601 | 0.786 | 0.790 | 0.852 | 1.316 | 1.316 |
| 2006 | 0.006 | 0.074 | 0.280 | 0.549 | 0.744 | 1.041 | 1.574 | 1.574 |
| 2007 | 0.007 | 0.101 | 0.484 | 1.026 | 1.358 | 1.744 | 2.126 | 2.126 |
| 2008 | 0.004 | 0.058 | 0.276 | 0.626 | 0.841 | 1.034 | 1.246 | 1.246 |
| 2009 | 0.003 | 0.045 | 0.220 | 0.528 | 0.785 | 1.001 | 1.175 | 1.175 |
| 2010 | 0.002 | 0.038 | 0.176 | 0.436 | 0.673 | 0.849 | 1.008 | 1.008 |
| 2011 | 0.002 | 0.033 | 0.141 | 0.324 | 0.540 | 0.719 | 0.881 | 0.881 |
| 2012 | 0.002 | 0.037 | 0.153 | 0.298 | 0.480 | 0.675 | 0.832 | 0.832 |
| 2013 | 0.002 | 0.037 | 0.158 | 0.300 | 0.498 | 0.713 | 0.876 | 0.876 |
| 2014 | 0.002 | 0.039 | 0.170 | 0.331 | 0.563 | 0.766 | 0.953 | 0.953 |
| 2015 | 0.003 | 0.057 | 0.249 | 0.435 | 0.660 | 0.877 | 1.125 | 1.125 |
| 2016 | 0.004 | 0.076 | 0.345 | 0.631 | 0.864 | 1.105 | 1.393 | 1.393 |
| 2017 | 0.004 | 0.072 | 0.351 | 0.636 | 0.884 | 1.113 | 1.413 | 1.413 |
| 2018 | 0.003 | 0.072 | 0.383 | 0.755 | 1.060 | 1.348 | 1.680 | 1.680 |
| 2019 | 0.002 | 0.048 | 0.266 | 0.656 | 1.013 | 1.279 | 1.556 | 1.556 |
| 2020 | 0.002 | 0.042 | 0.225 | 0.576 | 0.915 | 1.240 | 1.486 | 1.486 |
| 2021 | 0.001 | 0.032 | 0.174 | 0.459 | 0.701 | 0.911 | 1.085 | 1.085 |

Table 11. Whiting in divisions 7.b, c, e-k. Stock number-at-age ('000).

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 2027897 | 291554 | 94860 | 37192 | 17484 | 14771 | 9914 | 2519 |
| 2000 | 1701339 | 606549 | 111836 | 31802 | 11652 | 4466 | 2993 | 1793 |
| 2001 | 1421549 | 498264 | 238876 | 39555 | 9914 | 2851 | 888 | 573 |
| 2002 | 1267864 | 412764 | 193577 | 90478 | 13543 | 2194 | 392 | 96 |
| 2003 | 998197 | 375735 | 162050 | 77773 | 38110 | 4332 | 507 | 48 |
| 2004 | 871785 | 292860 | 145356 | 62250 | 36758 | 13870 | 1025 | 81 |
| 2005 | 749519 | 244099 | 106380 | 40407 | 18132 | 13568 | 3999 | 206 |
| 2006 | 826353 | 219843 | 81238 | 29472 | 10970 | 5145 | 4453 | 775 |
| 2007 | 1020649 | 236347 | 86603 | 35050 | 10171 | 3388 | 1172 | 750 |
| 2008 | 1218297 | 300794 | 86509 | 27784 | 7770 | 1685 | 387 | 158 |
| 2009 | 1278262 | 363220 | 117262 | 33160 | 8845 | 2207 | 408 | 108 |
| 2010 | 652585 | 395419 | 141166 | 48649 | 11743 | 2594 | 539 | 110 |
| 2011 | 624387 | 192617 | 166946 | 59400 | 18401 | 3859 | 746 | 163 |
| 2012 | 579921 | 185721 | 81198 | 78604 | 25616 | 6696 | 1261 | 259 |
| 2013 | 1205680 | 160105 | 76808 | 36487 | 37398 | 10450 | 2232 | 455 |
| 2014 | 545184 | 391306 | 64226 | 34708 | 16660 | 15588 | 3404 | 764 |
| 2015 | 507319 | 160542 | 175051 | 28967 | 15365 | 6018 | 4993 | 1104 |
| 2016 | 330789 | 148074 | 64806 | 75519 | 11911 | 5242 | 1671 | 1364 |
| 2017 | 293206 | 91972 | 55229 | 22804 | 24183 | 3382 | 1167 | 519 |
| 2018 | 504093 | 80182 | 34322 | 19539 | 7271 | 6424 | 760 | 283 |
| 2019 | 400108 | 152847 | 30350 | 11503 | 5484 | 1658 | 1110 | 133 |
| 2020 | 494034 | 117017 | 60296 | 11918 | 3418 | 1255 | 313 | 180 |
| 2021 | 353025 | 149859 | 49413 | 25825 | 4057 | 875 | 237 | 77 |

## Table 12. Whiting in divisions 7.b, c, e-k. Summary table.

| Year | Recruitment age 0 |  |  | SSB |  |  | Landings | Discards | F ages 2-5 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Value | Low | High | Value | Low | High |  |  | Value | Low | High |
| 1999 | 2027897 | 3019160 | 1362089 | 54302 | 61100 | 48260 | 20180 | 5420 | 0.80 | 0.94 | 0.69 |
| 2000 | 1701339 | 2486760 | 1163986 | 52501 | 60164 | 45815 | 15644 | 4400 | 0.81 | 0.94 | 0.69 |
| 2001 | 1421549 | 1930257 | 1046908 | 61432 | 73583 | 51288 | 13196 | 9877 | 0.89 | 1.03 | 0.76 |
| 2002 | 1267864 | 1703340 | 943722 | 70799 | 84286 | 59469 | 13640 | 7336 | 0.62 | 0.75 | 0.52 |
| 2003 | 998197 | 1346874 | 739785 | 68985 | 78939 | 60286 | 11788 | 10337 | 0.57 | 0.69 | 0.47 |
| 2004 | 871785 | 1180819 | 643630 | 65094 | 73768 | 57440 | 10321 | 19522 | 0.72 | 0.85 | 0.61 |
| 2005 | 749519 | 1007845 | 557406 | 46453 | 52199 | 41340 | 12575 | 13598 | 0.76 | 0.88 | 0.65 |
| 2006 | 826353 | 1135740 | 601247 | 41147 | 46292 | 36574 | 9908 | 5098 | 0.65 | 0.79 | 0.55 |
| 2007 | 1020649 | 1374910 | 757667 | 39139 | 44757 | 34226 | 9424 | 8439 | 1.15 | 1.34 | 0.99 |
| 2008 | 1218297 | 1640683 | 904652 | 37123 | 42658 | 32306 | 6080 | 3760 | 0.69 | 0.83 | 0.58 |
| 2009 | 1278262 | 1870293 | 873636 | 46729 | 53865 | 40539 | 6574 | 4281 | 0.63 | 0.77 | 0.52 |
| 2010 | 652585 | 881377 | 483184 | 68742 | 79815 | 59205 | 9570 | 5346 | 0.53 | 0.66 | 0.43 |
| 2011 | 624387 | 836391 | 466120 | 67898 | 78585 | 58664 | 10084 | 3750 | 0.43 | 0.54 | 0.34 |
| 2012 | 579921 | 788972 | 426262 | 61381 | 70051 | 53784 | 10834 | 5116 | 0.40 | 0.50 | 0.32 |
| 2013 | 1205680 | 1940367 | 749170 | 55194 | 62296 | 48902 | 12131 | 4026 | 0.42 | 0.52 | 0.34 |


| Year | Recruitment age 0 |  |  | SSB |  |  | Landings | Discards | F ages 2-5 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Value | Low | High | Value | Low | High |  |  | Value | Low | High |
| 2014 | 545184 | 730884 | 406667 | 60670 | 69406 | 53033 | 12983 | 4672 | 0.46 | 0.56 | 0.38 |
| 2015 | 507319 | 692155 | 371843 | 61195 | 69360 | 53992 | 13110 | 6528 | 0.56 | 0.66 | 0.47 |
| 2016 | 330789 | 493952 | 221522 | 49278 | 56528 | 42957 | 15201 | 8259 | 0.74 | 0.87 | 0.63 |
| 2017 | 293206 | 434204 | 197994 | 33504 | 37662 | 29806 | 12377 | 2791 | 0.75 | 0.87 | 0.64 |
| 2018 | 504093 | 707837 | 358994 | 23130 | 26061 | 20529 | 9007 | 2139 | 0.89 | 1.04 | 0.75 |
| 2019 | 400108 | 571894 | 279923 | 22862 | 27150 | 19251 | 6588 | 970 | 0.80 | 0.99 | 0.65 |
| 2020 | 494034 | 781842 | 312172 | 27954 | 34842 | 22428 | 5931 | 1266 | 0.74 | 1.06 | 0.51 |
| 2021 | 353025 | 731427 | 170388 | 26600 | 35064 | 20180 | 6153 | 1224 | 0.56 | 1.01 | 0.31 |
| 2022 | 400108* | 507319 | 293206 | 32346 | 46048 | 22581 |  |  |  |  |  |

* Median resampled (2015-2021).

Table 13. Whiting in divisions 7.b, c, e-k. Management options table.

| Basis | Total catch (2023) | Projected landings (2023) | Projected discards (2023) | $F_{\text {total }}$ <br> (2023) | $F_{\text {projected land- }}$ ings (2023) | $F_{\text {projected dis- }}$ cards (2023) | $\begin{aligned} & \text { SSB } \\ & \text { (2024) } \end{aligned}$ | $\begin{aligned} & \text { \% SSB } \\ & \text { change* } \end{aligned}$ | \% advice change ** | \% Probability of being below $\mathrm{B}_{\text {lim }}$ in 2024 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |  |  |
| MSY and precautionary considerations: $\mathrm{F}=0$ | 0 | 0 | 0 | 0.000 | 0.000 | 0.000 | 38109 | 26 | -100 | 40 |
| Other options |  |  |  |  |  |  |  |  |  |  |
| MSY approach: | 4030 | 3542 | 488 | 0.224 | 0.199 | 0.025 | 34568 | 14 | -9.5 | 62 |
| $\mathrm{F}_{\text {MSY }} \times$ SSB $_{2023} / \mathrm{MSY}^{\text {B }}$ trigger ${ }^{\text {a }}$ |  |  |  |  |  |  |  |  |  |  |
| EU MAP^^: | 4030 | 3542 | 488 | 0.224 | 0.199 | 0.025 | 34568 | 14 | -9.5 | 62 |
| $\mathrm{F}_{\text {MSY }} \times$ SSB $_{2023} /$ MSY $_{\text {trigger }}$ |  |  |  |  |  |  |  |  |  |  |
| EU MAP^^: | 3441 | 3026 | 415 | 0.188 | 0.167 | 0.021 | 35065 | 16 | -23 | 59 |
| $\mathrm{F}_{\text {MSY lower }} \times$ SSB ${ }_{\text {2023 }} /$ MSY $^{\text {Etrigger }}$ |  |  |  |  |  |  |  |  |  |  |
| $F=F_{M S Y}=F_{\text {pa }}$ | 6322 | 5532 | 790 | 0.375 | 0.333 | 0.042 | 32641 | 8 | 42 | 72 |
| $\mathrm{SSB}_{2024}=\mathrm{Bl}_{\text {lim }}$ | 1715 | 1512 | 203 | 0.090 | 0.079 | 0.011 | 36571 | 21 | -61 | 50 |
| $\mathrm{SSB}_{2024}=\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\text {trigger }} \wedge$ |  |  |  |  |  |  |  |  |  |  |
| $F=F_{2022}$ | 10376 | 9013 | 1363 | 0.701 | 0.622 | 0.079 | 29323 | -3 | 133 | 86 |
| $\mathrm{SSB}_{2024}=\mathrm{SSB}_{2023}$ | 9108 | 7939 | 1169 | 0.590 | 0.523 | 0.067 | 30343 | 0 | 105 | 82 |

Input units are thousands and kg , outputs in tonnes.

Table 14. Whiting in divisions 7.b, c, e-k. Basis for the catch forecast scenarios.

| Variable | Value | Notes |
| :--- | :---: | :--- |
| Fages 2-5 (2022) | 0.701 | F based on catch of 9240 tonnes for 2022 |
| SSB (2023) | 30343 | Short-term forecast fishing at F = 0.571; in tonnes |
| Recruitment age 0 (2022- <br> 2023) | 400108 | Median resampled (2015-2021); in thousands |
| Catch (2022) | 10741 | Catch based on mixed fisheries considerations (ICES, 2021a) when <br> haddock is fished in 2022 at $\mathrm{F}=0.582 ;$ in tonnes |
| Projected landings (2022) | 9228 | Short-term forecast assuming average 2019-2021 landings pattern; in <br> tonnes |
| Projected discards (2022) | 1513 | Short-term forecast assuming average 2019-2021 discard pattern; in <br> tonnes |

Input units are thousands and kg output in tonnes.

### 38.13 Figures




Figure 1. Distribution of international OTB effort within the Irish EEZ 2014-2018.


Figure 2. Whiting in 7.b-ce-k (Celtic Sea), annual Landings (Green) and Discards- (red) at-age.


Figure 3. Whiting in 7.b-ce-k (Celtic Sea), annual Landings- (Blue) and Discards- (red) at-age. Dashed lines give revised data uploaded to InterCatch. Solid lines show the final raised International Catch Numbers-at-Age used in the assessment.

Whiting.27.7b-ce-k Rivard Corrected stock weights


Figure 4. Whiting in 7.b, c, e-k (Celtic Sea). Rivard corrected stock weights-at-age.


Figure 5. Annual proportions of Discarding (by weight) for the Celtic Sea whiting revised time-series (2003-2021).


Figure 6. Proportion of landings (upper panels) and discards (lower panels) for Celtic Sea whiting (2003-2021).

VAST index internal consistency


Figure 7. Whiting in 7.b, c, e-k (Celtic Sea). Pairwise scatterplots for the log numbers-at-age for the VAST combined survey index.


Figure 8. Whiting in 7.e-k (Celtic Sea). Mean log standardized plots of combined IE-IGFS \& FR-EVHOE indices by year class (top panel) and by year (lower panel). Only age 0-2 is included in the assessment.


Figure 9. Whiting in 7.b, c, e-k (Celtic Sea). SAM assessment summary plots of SSB (top left), Fbar2-5 (top right) and recruitment (bottom left) at Age $\mathbf{0}$. Grey line and shaded area indicate the previous year's assessment. An overall downward trend in biomass and recruitment since the last small pulse in 2013 is evident and followed by a significant drop in fishing effort as that biomass was removed.


Figure 10. Fit to the catch-at-age data (top) and VAST index (bottom left) for final SAM assessment run. Model fits for commercial biomass index are given in lower right panel. Point observations are presented along with model prediction lines.


Figure 11. Residual patterns for the catch at age data (top), commercial biomass index (middle) and VAST IBTS index (bottom) for final SAM assessment run.


Figure 12. Retrospective patters and Mohn's Rho calculations for SSB (top left), $\mathrm{F}_{\text {bar }}$ (top right) and recruitment (bottom left).


Figure 13. Contribution to advised catch and Spawning-Stock Biomass (SSB) of the recruitment assumption used in the short term forecast.

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## Annex 2: Stock Annexes

The table below provides an overview of the WGCSE Stock Annexes. Stock Annexes for other stocks are available on the ICES website Library under the Publication Type "Stock Annexes". Use the search facility to find a particular Stock Annex, refining your search in the left-hand column to include the year, ecoregion, species, and acronym of the relevant ICES expert group.

| Stock ID | Stock name | Last updated | Link |
| :---: | :---: | :---: | :---: |
| anf.27.3a46 | Anglerfish (Lophius budegassa, Lophius piscatorius) in subareas 4 and 6, and in Division 3.a (North Sea, Rockall and West of Scotland, Skagerrak and Kattegat) | October 2019 | Anglerfish 3.a46 |
| bss.27.4bc7d-h | Seabass (Dicentrarchus labrax) in divisions 4.b-c, 7.a, and 7.d-h (central and southern North Sea, Irish Sea, English Channel, Bristol Channel, and Celtic Sea) | May 2020 | Sea bass 47 |
| cod.27.7e-k | Cod (Gadus morhua) in divisions 7.e-k (eastern English Channel and southern Celtic Seas) | October 2020 | Cod 7.e-k |
| cod.27.7a | Cod (Gadus morhua) in Division 7.a (Irish Sea) | February 2022 | Cod 7.a |
| cod.27.6b | Cod (Gadus morhua) in Division 6.b (Rockall) | May 2013 | Cod 6.b |
| cod.27.6a | Cod (Gadus morhua) in Division 6.a (West of Scotland) | May 2022 | Cod $6 . a$ |
| gug-celt | Grey gurnard in Subarea 6 and Divisions 7.a-c and e-k | March 2014 | Grey gurnard |
| had.27.7b-k | Haddock (Melanogrammus aeglefinus) in divisions 7.b-k (southern Celtic Seas and English Channel) | May 2017 | Haddock 7.b-k |
| had.27.7a | Haddock (Melanogrammus aeglefinus) in Division 7.a (Irish Sea) | June 2021 | Haddock 7.a |
| had.27.6b | Haddock (Melanogrammus aeglefinus) in Division 6.b (Rockall) | May 2020 | Haddock 6.b |
| lez.27.4a6a | Megrim (Lepidorhombus ssp.) in divisions 4.a and 6.a (northern North Sea, West of Scotland) | June 2021 | Megrim 4a6a |
| nep.fu. 11 | Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 11 (West of Scotland, North Minch) | May 2016 | Nephrops FU11 |
| nep.fu. 12 | Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 12 (West of Scotland, South Minch) | May 2016 | Nephrops FU12 |


| Stock ID | Stock name | Last updated | Link |
| :---: | :---: | :---: | :---: |
| nep.fu. 13 | Norway lobster (Nephrops norvegicus) in Division 6.a, Functional Unit 13 (West of Scotland, the Firth of Clyde and Sound of Jura) | May 2017 | Nephrops FU13 |
| nep.fu. 14 | Norway lobster (Nephrops norvegicus) in Division 7.a, Functional Unit 14 (Irish Sea, East) | $\begin{aligned} & \text { September } \\ & 2018 \end{aligned}$ | Nephrops FU14 |
| nep.fu. 15 | Norway lobster (Nephrops norvegicus) in Division 7.a, Functional Unit 15 (Irish Sea, West) | May 2018 | Nephrops FU15 |
| nep.fu. 16 | Norway lobster (Nephrops norvegicus) in divisions 7.b-c and 7.j-k, Functional Unit 16 (west and southwest of Ireland, Porcupine Bank) | March 2013 | Nephrops FU16 |
| nep.fu. 17 | Norway lobster (Nephrops norvegicus) in Division 7.b, Functional Unit 17 (west of Ireland, Aran grounds) | May 2016 | Nephrops FU17 |
| nep.fu. 19 | Norway lobster (Nephrops norvegicus) in divisions 7.a, 7.g, and 7.j, Functional Unit 19 (Irish Sea, Celtic Sea, eastern part of southwest of Ireland) | October 2019 | Nephrops FU19 |
| nep.fu. 2021 | Norway lobster (Nephrops norvegicus) in divisions 7.g and 7.h, functional units 20 and 21 (Celtic Sea) | October 2019 | Nephrops FU2021 |
| nep.fu. 22 | Norway lobster (Nephrops norvegicus) in divisions 7.g and 7.f, Functional Unit 22 (Celtic Sea, Bristol Channel) | May 2018 | Nephrops FU22 |
| nep.fu. 2324 | Norway lobster (Nephrops norvegicus) in divisions 8.a and 8.b, functional units 23-24 (northern and central Bay of Biscay) |  | Not available |
| ple.27.7bc | Plaice (Pleuronectes platessa) in divisions 7.b-c (West of Ireland) | April 2013 | Plaice 7.bc |
| ple.27.7h-k | Plaice (Pleuronectes platessa) in divisions 7h-k (Celtic Sea South, southwest of Ireland) | May 2021 | Plaice 7.h-k |
| ple.27.7fg | Plaice (Pleuronectes platessa) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) | May 2022 | Plaice 7.fg |
| ple.27.7e | Plaice (Pleuronectes platessa) in Division 7.e (western English Channel) | May 2022 | Plaice 7.e |
| ple.27.7a | Plaice (Pleuronectes platessa) in Division 7.a (Irish Sea) | June 2021 | Plaice 7.a |
| sol.27.7bc | Sole (Solea solea) in divisions 7.b and 7.c (West of Ireland) | April 2013 | Sole 7.bc |
| sol.27.7h-k | Sole (Solea solea) in divisions 7.h-k (Celtic Sea South, Southwest of Ireland) | May 2020 | Sole 7.h-k |
| sol.27.7fg | Sole (Solea solea) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) | May 2022 | Sole 7.fg |


| Stock ID | Stock name | Last updated | Link |
| :--- | :--- | :--- | :--- |
| sol.27.7e | Sole (Solea solea) in Division 7.e (western English Channel) | June 2021 | Sole 7.e |
| sol.27.7a | Sole (Solea solea) in Division 7.a (Irish Sea) | May 2022 | Sole 7.a |
| whg.27.7b-ce-k | Whiting (Merlangius merlangus) in divisions 7.b-c and 7.e-k (southern Celtic Seas and eastern English Channel) | September | Whiting 7.bc,e-k |
| whg.27.7a | Whiting (Merlangius merlangus) in Division 7.a (Irish Sea) | 2020 | May 2017 |
| whg.27.6b | Whiting (Merlangius merlangus) in Division 6.b (Rockall) | Whiting 7.a |  |
| whg.27.6a | Whiting (Merlangius merlangus) in Division 6.a (West of Scotland) | May 2013 | Whiting 6.b |

## Annex 3: Working documents

The following working documents were presented to WGCSE in spring 2022. They are found in full in the following pages:
WD1: Maturity-at-age estimates for Irish Demersal Stocks in 6.a and 7.b-k between 2004-2021. Sara-Jane Moore and Hans Gerritsen.

WD2: Reducing uncertainty and assessing Bias in estimates of Nephrops norvegicus population size. Niall G. Fallon.

WD3: Development of Reference Points for the assessment of Celtic Sea Plaice Divisions 7.f and 7.g. Timothy Earl, Vladimir Laptikhovsky, Mathieu Lundy, Jonathan White.

# Maturity-at-age estimates for Irish Demersal Stocks in 6.a and 7.b-k between 2004-2021 

Sara-Jane Moore and Hans Gerritsen<br>Marine Institute<br>Galway<br>Ireland

## Introduction

This document provides maturity-at-age estimates for stocks assessed by the WGCSE and WGBIE. All data are obtained on surveys and commercial sampling carried out by the Marine Institute.

## Methods

Data was used from the Marine Institute Q1 Biological sampling programme (2010-2021), AtSea Observer programme (2010-2021), Irish Anglerfish and megrim survey (2016-2021), the Irish beam trawl Ecosystem survey (2016-2018) and the MI Biological sampling survey (20042009). Sampling levels were reduced in 2021 as a result of COVID 19 and diminished access to at-sea samples and also port samples. Proportions mature-at-age were estimated by constructing a matrix containing the sample numbers by age, sex and maturity state (mature/immature) at each length class. Unsexed individuals (usually small fish with undeveloped gonads) were assigned in equal numbers to both sexes. This Age-Sex-MaturityLength Key (ASMLK) was applied to the length-frequency data to estimate the proportions mature-at-age for either sex and both sexes combined. Any gaps in the ASMLK were filled in using a multinomial model (Gerritsen et al., 2006).

## Results

Because overall there was no clear evidence of trends in maturity over time for any stock, data from all years (2004-2021) were combined. Overall, the perception of age at maturity has not changed from previous years working documents. Figure 1 shows that for most stocks there are no clear trends in the $\mathrm{L}_{50}$ over time. Estimates for cod in area $7 \mathrm{e}-\mathrm{k}$ varied from around 40 cm to 60 cm , however the sample sizes for this stock were generally very low at the start of the time-series. In the last 3 years, $\mathrm{L}_{50}$ estimates are consistent with an average of 48
cm . Sole in area 7 shows variable estimates in recent years. Plaice in area 7 shows two outlying estimates in 2013 and 2019 but these was estimated with low precision. There is a slight decrease $\mathrm{L}_{50}$ estimates from 30 cm in 2020 to 24 cm in 2021. Whiting in $7 \mathrm{~b}-\mathrm{k}$ shows a decline in $\mathrm{L}_{50}$ in 2019 but the data is based on low sample levels. In 2020 and 2021 the $\mathrm{L}_{50}$ increased to an average of 21 cm .

Table 1(a) shows the estimated proportions mature-at-age. "All" sexes is a weighted maturity ogive and included unsexed individuals most likely to be immature. For Megrim 7\&8, maturity at age is slightly lower than that used by WGBIE. Estimated proportions mature for plaice and sole were also slightly lower than those used by the working group. Whiting and Haddock estimates for 6a are variable over time.

Discussion
Some (relatively minor) differences were found between the ogives used by the working groups and the current findings. Because Irish sampling generally does not cover the full extent of the stocks, it is difficult to determine whether the Irish estimates are unbiased. It is possible that the lack of full spatial coverage can explain some of the differences.

## References

Gerritsen, H.D., Armstrong, M.J., Allen, M., McCurdy, W.J. and Peel, J.A.D., 2003. Variability in maturity and growth in a heavily exploited stock: whiting (Merlangius merlangus L.) in the Irish Sea. J. Sea Res., 49(1): 69-82.
Gerritsen, H.D., McGrath, D. and Lordan, C., 2006. A simple method for comparing age-length keys reveals significant regional differences within a single stock of haddock (Melanogrammus aeglefinus). ICES J. Mar. Sci., 63(3): 1096-1100.
ICES, 2018. Report of the Working Group for the Celtic Seas Ecoregion (WGCSE), 9-18 May, 2018, ICES CM 2018/ACOM:13
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Figure 1. Length at $50 \%$ maturity (L50; cm) for females by stock and year.

Table 1 (a). Estimated proportions mature (sample numbers in table below) by stock, sex and age. Maturity ogives used by the WG are also given.

| Stock | Sex/Wg | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| cod-7e-k | All | 0 | 0.39 | 0.83 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |
|  | F | 0.02 | 0.47 | 0.91 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |
|  | M | 0.01 | 0.57 | 0.97 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | WKCELTIC | 0 | 0.54 | 0.93 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |
| had-7b-k | All | 0.21 | 0.83 | 0.93 | 0.95 | 0.97 | 0.96 | 0.98 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |
|  | F | 0.01 | 0.91 | 0.97 | 0.98 | 0.97 | 0.99 | 0.98 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |
|  | M | 0.28 | 0.82 | 0.91 | 0.91 | 1.0 | 0.92 | 0.95 | 1 | 1 |  |  | 1 |  |  |  |  |  |  |
|  | WKCELTIC | 0.04 | 0.91 | 0.97 | 0.98 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |
| had 6.a | All | 0.07 | 0.56 | 0.57 | 0.73 | 0.69 | 0.77 | 0.75 | 0.86 | 0.81 | 1 |  |  |  |  |  |  |  |  |
|  | F | 0.03 | 0.78 | 0.75 | 0.81 | 0.85 | 0.89 | 0.80 | 1 | 0.90 | 1 |  |  |  |  |  |  |  |  |
|  | M | 0.10 | 0.65 | 0.59 | 0.70 | 0.45 | 0.40 | 0.68 | 0.33 | 0.42 |  |  |  |  |  |  |  |  |  |
|  | WGNSSK | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |
| mgw-78 | All | 0.30 | 0.31 | 0.57 | 0.74 | 0.75 | 0.76 | 0.76 | 0.84 | 0.85 | 0.86 | 0.90 | 0.88 | 0.93 | 0.92 | 0.93 | 0.94 | 1 | 1 |
|  | F | 0.11 | 0.29 | 0.65 | 0.86 | 0.86 | 0.86 | 0.81 | 0.82 | 0.86 | 0.87 | 0.88 | 0.86 | 0.92 | 0.90 | 0.93 | 0.95 | 1 | 1 |
|  | M | 0.62 | 0.35 | 0.54 | 0.69 | 0.72 | 0.77 | 0.84 | 0.88 | 0.87 | 0.82 | 0.84 | 0.93 | 1 | 1 | 1 | 1 |  |  |
|  | WGBIE | 0.04 | 0.21 | 0.6 | 0.9 | 0.98 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |
| ple-7 | All | 0 | 0.27 | 0.50 | 0.66 | 0.76 | 0.87 | 0.87 | 0.87 | 0.89 | 0.95 | 0.68 | 0.87 |  |  |  |  |  |  |
|  | F | 0 | 0.14 | 0.44 | 0.65 | 0.77 | 0.89 | 0.89 | 0.83 | 0.83 | 0.91 | 0.86 | 0.78 |  |  |  |  |  |  |
|  | M | 0 | 0.33 | 0.59 | 0.73 | 0.82 | 0.88 | 0.84 | 0.92 | 0.86 | 1 | 1 | 1 |  |  |  |  |  |  |
| ple 7.a | WGCSE | 0 | 0.24 | 0.57 | 0.74 | 0.93 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |
| ple 7.fg | WGCSE | 0 | 0.26 | 0.52 | 0.86 | 1 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |
| sol-7 | All | 0 | 0.15 | 0.39 | 0.52 | 0.79 | 0.79 | 0.83 | 0.73 | 0.82 | 0.89 | 0.89 | 0.89 | 0.93 | 0.88 | 0.75 | 0.97 | 0.97 | 0.98 |
|  | F | 0 | 0.14 | 0.46 | 0.63 | 0.85 | 0.92 | 0.95 | 0.97 | 0.95 | 0.96 | 0.95 | 0.95 | 0.94 | 1 | 0.90 | 1 | 1 | 1 |
|  | M | 0 | 0.22 | 0.40 | 0.49 | 0.56 | 0.71 | 0.70 | 0.75 | 0.68 | 0.71 | 0.86 | 0.73 | 0.75 | 0.67 | 0.52 | 0.78 | 0.68 | 0.70 |
| sol 7.fg | WGCSE | 0 | 0.14 | 0.45 | 0.88 | 0.98 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |
| whg 7b-k | All | 0.50 | 0.90 | 0.96 | 0.91 | 0.87 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | F | 0.29 | 0.96 | 0.98 | 0.98 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | M | 0.49 | 0.83 | 0.95 | 0.85 | 0.80 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | WKCELTIC | 0.61 | 0.94 | 0.97 | 0.97 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |
| whg 6.a | All | 0.40 | 0.70 | 0.65 | 0.80 | 0.88 | 0.81 | 0.93 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | F | 0.43 | 0.88 | 0.93 | 0.92 | 0.96 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | M | 0.47 | 0.59 | 0.50 | 0.64 | 0.58 | 0.38 | 0.78 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | WGCSE | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |

Table 1 (b). Sample numbers by stock, sex and age for associated maturity in Table 1(a) above.

| Stock | Sex/Wg | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| cod-7e-k | All | 1646 | 1881 | 172 | 25 | 2 | 2 | 3 |  |  |  |  |  |  |  |  |  |  |  |
|  | F | 734 | 849 | 82 | 14 | 2 | 2 | 3 |  |  |  |  |  |  |  |  |  |  |  |
|  | M | 837 | 1015 | 89 | 11 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| had-7b-k | All | 1228 | 2172 | 1603 | 540 | 269 | 167 | 105 | 44 | 17 | 4 | 1 | 1 |  |  |  |  |  |  |
|  | F | 490 | 1220 | 947 | 320 | 167 | 113 | 70 | 27 | 13 | 4 | 1 |  |  |  |  |  |  |  |
|  | M | 579 | 944 | 654 | 220 | 101 | 54 | 35 | 17 | 4 |  |  | 1 |  |  |  |  |  |  |
| had-scow | All | 143 | 535 | 499 | 310 | 291 | 193 | 130 | 48 | 32 | 8 |  |  |  |  |  |  |  |  |
|  | F | 58 | 281 | 305 | 207 | 202 | 140 | 95 | 37 | 25 | 7 |  |  |  |  |  |  |  |  |
|  | M | 66 | 250 | 194 | 103 | 89 | 53 | 35 | 11 | 7 |  |  |  |  |  |  |  |  |  |
| mgw-78 | All | 57 | 1268 | 2114 | 1571 | 970 | 672 | 413 | 260 | 191 | 156 | 140 | 114 | 60 | 53 | 44 | 30 | 24 | 9 |
|  | F | 34 | 622 | 1301 | 1081 | 699 | 463 | 257 | 178 | 153 | 133 | 135 | 107 | 58 | 53 | 44 | 29 | 24 | 9 |
|  | M | 22 | 640 | 801 | 489 | 271 | 209 | 156 | 82 | 38 | 23 | 5 | 7 | 2 |  |  | 1 |  |  |
| ple-7 | All | 28 | 525 | 1375 | 1058 | 730 | 348 | 192 | 141 | 57 | 33 | 21 | 9 |  |  |  |  |  |  |
|  | F | 13 | 233 | 793 | 628 | 491 | 230 | 131 | 94 | 42 | 25 | 19 | 8 |  |  |  |  |  |  |
|  | M | 14 | 267 | 556 | 417 | 233 | 118 | 60 | 46 | 15 | 8 | 2 | 1 |  |  |  |  |  |  |
| sol-7 | All | 2 | 75 | 649 | 972 | 693 | 472 | 333 | 235 | 168 | 93 | 70 | 62 | 52 | 28 | 14 | 12 | 12 | 8 |
|  | F | 2 | 46 | 537 | 823 | 584 | 327 | 202 | 114 | 90 | 45 | 24 | 28 | 27 | 12 | 8 | 8 | 5 | 5 |
|  | M |  | 24 | 111 | 148 | 108 | 144 | 130 | 121 | 77 | 46 | 45 | 34 | 25 | 16 | 5 | 4 | 7 | 3 |
| sol 7.fg |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| whg-7b-k | All | 1300 | 1360 | 904 | 381 | 127 | 28 | 4 | 2 |  |  |  |  |  |  |  |  |  |  |
|  | F | 601 | 773 | 510 | 212 | 70 | 12 | 2 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | M | 642 | 584 | 394 | 169 | 56 | 16 | 2 | 1 |  |  |  |  |  |  |  |  |  |  |
| whg-scow | All | 200 | 416 | 472 | 281 | 153 | 59 | 17 | 3 |  |  |  |  |  |  |  |  |  |  |
|  | F | 95 | 224 | 265 | 177 | 112 | 38 | 7 | 2 |  |  |  |  |  |  |  |  |  |  |
|  | M | 103 | 192 | 207 | 104 | 41 | 21 | 10 | 1 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

# Working Paper: Reducing uncertainty \& assessing Bias in estimates of Nephrops norvegicus population size 

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Introduction

The fishery for Nephrops norvegicus is one of the most valuable in Scotland ( $£ 86 \mathrm{~m}$ in 2019), representing ${ }^{\sim} 15 \%$ of the value of all vessels' landings. The South Minch Nephrops Functional Unit (FU 12), found off the northwest coast of Scotland (Fig. 1, inset) has been surveyed on an annual basis since 1995. The total abundance of Nephrops is estimated using burrow density data collected during a research vessel based underwater television survey (UWTV) (ICES, 2021). Estimates of stock size derived from UWTV survey sample data are subject to uncertainty arising from measurement error (i.e. sampling uncertainty). Minimising measurement error is essential to the calculation of survey quantities with a level of uncertainty which allows for the evaluation of trends in stock dynamics.


Figure 1. FU 12 is divided into three areal strata (East, South, and West), and three sediment strata (Mud green, Sandy Mud blue, and Muddy sand yellow), for UWTV sample allocation purposes.

The current sampling scheme for FU 12 is stratified across three sediment types (muddy sand, sandy mud and mud, following the Folk sediment classification) with fixed proportions of sampling effort in each of three areas (East, South and West; Fig. 1). Data from UWTV surveys of FU 12 are characterised by relatively high sample variance due to variability in burrow density within and between sediment types, when compared with the other Scottish Nephrops functional units. Estimates of abundance derived from FU 12 survey data using the current method ("standard method", "standard abundance estimates"), a stratified
mean estimator, therefore have relatively high uncertainty, affecting the precise detection of temporal trends in abundance. In addition, until 2021 UWTV survey samples were not allocated in direct proportion to area in the case of two strata, Eastern Sandy Mud and Western Muddy Sand, and samples were not always collected in the Western Mud stratum, along with some other less substantial temporal inconsistencies (Fig. 2). Burrow counts in the Eastern Sandy Mud stratum are notable for being relatively high (between $\sim 0.6$ and 1 burrow per $\mathrm{m}^{2}$ ), particularly in recent years, and thus their over-representation in the sample set may have been problematic. Although the disparities in proportionate allocations are seemingly low (<10\%), the sample allocation scheme could potentially have been introducing a bias to abundance estimates. In 2021, a sample allocation scheme was implemented in which sample numbers per stratum were as close to proportionate to sediment areal coverage as sample size and logistics would allow.


Figure 2. Proportionate sample allocations by stratum 2006-2021.

The aim of this study is to identify an abundance estimation method for FU 12 Nephrops which has lower uncertainty when compared to the standard method, and to determine whether bias was being introduced to abundance estimates by the UWTV survey sample allocation method. Kriging is used here to generate estimates of Nephrops abundance, and these estimates and associated uncertainty measures are compared to those derived using the standard method. In order to evaluate the bias (if any) in abundance estimates that may have been introduced due to disproportionate sample allocations, a
resampling routine is used to generate estimates based on the standard method, where samples are taken in direct proportion to the area of each survey stratum.

## Methods

Geostatistical analyses were implemented in RGeostats (MINES ParisTech / ARMINES, 2020) to generate spatially explicit kriged surfaces of FU 12 Nephrops burrow densities based on UWTV survey data from 2006-2021 (See Petitgas et al. 2017 for detailed descriptions of geostatistical methods), as well as globally kriged mean density estimates to be raised to total abundance. The first step in kriging involves characterising the spatial structure of the variable of interest (Nephrops burrow density) using variography (Rivoirard et al., 2000): i.e. the calculation and modelling of variability in density as a function of sample separation distance. Experimental variograms were calculated from the survey density data, to which exponential variogram models were fitted (Fig. 3).


Figure 3. Variogram calculated for FU 12 Nephrops, from 2021 UWTV survey data. The experimental variogram of burrow density is represented by the dashed orange line. The model variogram is represented by the solid orange line.

Once an acceptable variogram model was fitted, kriged surfaces of burrow densities across the estimation area were generated. Here, a grid size of $500 \times 500$ metres was used. A globally kriged mean abundance (numbers per $\mathrm{m}^{-2}$ ) was calculated, and raised to the total area of the main sediment patch (Fig. 4), with $95 \%$ confidence intervals calculated from the coefficient of variation. The abundance of Nephrops across the remaining, smaller sediment patches was calculated using the standard method, and added to the kriged total abundance. Sensitivity of estimates to grid size and variogram parameters were tested.


Figure 4. Sediment patches in FU 12 used in the calculation of Nephrops abundance. The main (blue) patch is the area across which kriging was carried out. Nephrops abundance across the remaining (orange) smaller patches was calculated using the standard method, and a total abundance was then summed across all sediment patches. The main sediment patch accounts for $82 \%$ of the survey domain by area.

In order to evaluate potential bias in UWTV survey standard abundance estimates, a resampling routine was implemented whereby a bootstrapped distribution of abundance estimates ( $n=1000$ ) was generated for each survey year using the standard method. Each estimate was derived based on a sample set which had a number of samples per stratum approximately proportionate to relative areal coverage of each stratum.

## Results

Kriging provides illustrative maps of the spatial distribution of Nephrops within the main sediment patch of FU 12 (e.g. Fig. 5).


Figure 5. Kriged Nephrops burrow density distribution for FU 12, calculated using 2011 UWTV burrow density data (overlaid as a black bubble plot where bubble area is proportional to burrow density, the black $\mathbf{x}$ symbols represent zero density observations). Darker red pixels represent areas of higher estimated Nephrops burrow density, and lighter yellow pixels represent areas of lower estimated density.

Kriged estimates of mean abundance were significantly lower than the estimates using the current method (Fig. $6 ; F_{(2,42)}=5.8, p<0.01$ ). The resampling-based mean abundance estimates tended to be closer to (but were also generally lower than) standard estimates, and followed the same overall temporal trend. There was no significant difference between the resample-based estimates and those derived using the standard method ( $F_{(2,42)}=5.8, p=0.36$ ). The 2021 resample-based estimate had the fourth lowest percentage deviation ( $1.54 \%$ ) from its equivalent standard estimate across the time series.


Figure 6. Time series estimates of FU 12 Nephrops abundance with $95 \%$ CIs using the standard method (blue line and polygon) overlaid with kriged estimates of mean total abundance with $95 \% \mathrm{Cls}$ (orange line and polygon), and resample estimates (black boxplots; the centre line is the mean, the whiskers are at the $95 \%$ quantiles of the bootstrap distribution).

Differences in magnitude aside, the three time series were highly correlated (pairwise Pearson correlation coefficients $>0.95$ ), suggesting generally good agreement between the trends observed across methods (Fig. 7). Compared to the standard method, there was a mean reduction of $34 \%$ in coefficient of variation using kriging (Fig. 8).


Figure 7. Pairwise comparisons of mean abundance estimates derived using kriging, the resampling routine, and the standard method. Each plot panel includes a red 1:1 line to aid in comparison of time series.


Figure 8. Time series of coefficients of variation for standard (blue line) and kriged (orange line) estimates of FU 12 Nephrops abundance.

Abundance estimates showed little sensitivity to changes in kriging grid size (Fig. 9) and variogram lag distances (Fig. 10).


Figure 9. FU 12 Nephrops kriged abundance estimates with $95 \%$ Cls calculated using a range of grid sizes.


Figure 10. FU 12 Nephrops kriged abundance estimates with $95 \%$ Cls calculated using a range of variogram lag distances.

Kriging can provide estimates of Nephrops abundance for FU 12 with reduced uncertainty when compared to the standard method, while being of a comparable magnitude and following similar historical trends. As such, kriging may offer a solution to the long-standing issue of highly uncertain abundance estimates for that management area. Ultimately, the outcome of the method relies heavily on the ability to fit a representative variogram model. It is thus important to fully explore the sensitivity of the variogram model fits to the assumptions applied in the calculation of the empirical variogram (e.g. distance lag). It may be useful to expand this analysis to multiple Scottish FUs to assess the performance of the kriging estimation method against the standard method in different scenarios.

Given the non-significant difference between the resampling-based method and the standard method, it does not appear that substantial bias has been introduced to the assessment due to the UWTV sample allocation method. Regardless, it would be favourable to correct the minor discrepancies in proportionate sample allocations for future surveys.

## References

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# Development of Reference Points for the assessment of Celtic Sea Plaice Divisions 7.f and 7.g 

v. 2.0 Post review update $\quad 20^{\text {th }}$ April, 2022

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## Introduction

The Benchmark meeting WKNSCS reviewed and developed a SAM assessment for Plaice in the Celtic Sea Divisions $7 \mathrm{f}-\mathrm{g}$. The assessment is accessible on Stockassessment.org as "plaice7fg 001". The details of the assessment are available in the WKNSCS report, section "Plaice (Pleuronectes platessa) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea)" under subsection "Suggested assessment"

Following review of the proposed assessment model during WKNSCS, reference points were developed (ICES 2021a). These were initially explored during WKNSCS and found to suggest a marked change in perception of stock status compared to the most recent qualitative assessment of the stock (ICES 2021b). A sub-group was tasked to review the proposed reference point formulation to understand the driver for the change.

The WKNSCS Benchmark meeting reviewed:
Benchmark Model 1 (BM1): a SAM model incorporating natural mortality estimates taken from the plaice assessment in the Irish Sea, Division 27.7a (ple.27.7a) (Stockassessment.org as "plaice7fg 001").

Subsequent to the WKNSCS Benchmark meeting, a second model and reference points derived from it were reviewed:

Benchmark Model 2 (BM2): A SAM model variant on BM1, using natural mortality estimates from the assessment of plaice in the English Channel (ple.27.7d), calculated using the Gislason equation, which were notably higher (Stockassessment.org as "Ple. 7 fg disc").

The two different estimates of natural mortality were used, as stock-specific data were considered not appropriate for the Celtic Sea ( 27.7 fg ) stock because they were lacking older ages. In both model variants, natural mortality was specified for each of the modelled age classes (1 to 10 year olds) and fixed across the full time series. Figure 1 compares the two different sets of natural mortality.


Figure 1. Comparison of natural mortalities of Benchmark Models 1 and 2.

The model variant focus, to develop reference points in this WD, is BM1, which is the model reviewed by the WKNSCS using natural mortalities taken from the plaice assessment in the Irish Sea, Division 27.7a.

## Method

Reference points were estimated following ICES Technical Guidelines (ICES, 2021) and WKMSYREF4 (ICES, 2017). In order to optimise for wanted catch in the reference point estimation, it was assumed that ages 1 and 2 have been unwanted and all other ages wanted, as age 2 corresponds with the Minimum Conservation Reference Size (MCRS) in this plaice stock.

Following review of an initial estimation of reference points, and following reviewer's recommendations, reference point estimation R -scripts were re-run in with the following settings:

Error settings were defined as:

```
Fbar (ple7fg)
sigma_m = sd(log(Fbar))
Fcv = standard deviation (Fbar) / mean (Fbar)
Fphi = 0.423 = default value (ICES, 2017)
SSBcv = 0.20 = default value (ICES, 2017)
```

Following further reviewer comments:

- $\mathrm{F}_{\text {MSY }}$ ranges were checked against $\mathrm{F}_{\mathrm{po5}}$, which resulted in no changes.
- The lower 5\%ile biomass was considered from fishing at $F_{\text {MSy }}$ for the second reference points (derived from BM2), but this was lower than $\mathrm{B}_{\mathrm{pa}}$, and so should not be used.


## Results

From plotting of Stock-Recruitment points, stock characteristics were taken as being of Type 5, "Stocks with no evidence that recruitment has been impaired or with no clear relation between stock and recruitment (no apparent S-R relationship)". For these stocks $B_{\text {loss }}$ is defined as a candidate value of $B_{\text {lim }}$, below which the dynamics of the stock are unknown. With $\mathrm{B}_{\text {loss }}$ taken from a stable part of the assessment and not from recent years. Blim then acts for the basis of estimation of "Precautionary Approach" and "MSY" reference points (ICES, 2021).

Reviewing the S-R data, the segmented regression inflection point is very close to this point and presents as a stable point from which to define $\mathrm{B}_{\text {loss, }}$ at 1344 tonnes. Following external review recommendation, the inflection point of the segmented regression was set to the $B_{\text {loss }}$ point. Appendix 1 sets out the RMarkdown of the described estimation.

## PA reference points:

$B_{\text {pa }}$ may be calculated as $=B_{\text {lim }} \times \exp (1.645 \times \sigma)$, with $\sigma$ estimated from the assessment uncertainty in SSB in the terminal year ( $\sigma$ is the estimated standard deviation of $\ln (S S B$ ) in the final assessment year). If $\sigma$ is unknown, 1.4 can be used as a default for " $\exp (1.645 \times \sigma$ )", equivalent to $\sigma=0.20$ ("SSBcv" in the above listed Error settings).

While $\mathrm{F}_{\text {lim }}$ may be estimated by simulating a stock with a segmented regression S-R relationship, with the point of inflection at $B_{\text {lim }}$ (as detailed above, set as $B_{\text {loss }}$ at 1344 tonnes) thus determining the $F=F_{\text {lim }}$ which, at equilibrium, yields a $50 \%$ probability of $S S B>B_{\text {lim }}$. This is may be done by running EqSIM to implement this simulation (ICES, 2021). $\mathrm{F}_{\text {lim }}$ is calculated with Fcv and Fphi both set to zero (ICES 2021).

## MSY reference points:

To define $F_{\text {MSy }}$ it is necessary to determine yield from the fishery, in the case of this plaice stock this is defined as the catch above Minimum Conservation Reference Size, and taken as wanted catch, aged above 2 year olds. Following the procedures set out in ICES (2021), FMSY was set to the median yield curve versus
$F$ : the value of $F$ that maximises yield. MSY $B_{\text {trigger, }}$ is set to safeguard against undesirable or unexpected low SSB when fishing at $F_{\text {MSY, }}$ as the $5^{\text {th }}$ percentile on the distribution of SSB when fishing at $\mathrm{F}_{\text {MSY }}$.

Following this process, through application of the R package and process "EqSIM" (set out in Appendix 1), derives reference points in Table 1. These, relative to the stock status (SSB) and fishing pressure (F) from the SAM assessment over time are shown in Figure 2 and Figure 3 respectively.

Table 1. BM1 - Estimated reference points for plaice 7.fg.

| Reference Point | Value | Rationale |
| :--- | :--- | :--- |
| MSY Btrigger | 1882 t | Bpa |
| Fmsy | 0.147 | Stochastic simulations with segmented regression \& Beverton-Holt |
| FmsyLower | 0.099 | Median lower point estimates of Stochastic simulations |
| FmsyUpper | 0.230 | Median upper point estimates of Stochastic simulations |
| Blim | 1344 t | Lowest observed SSB |
| Bpa | 1882 t | Blim combined with the assessment error |
| Flim | 1.275 | F with 50\% probability of SSB less than Blim |
| Fpa | 0.592 | F with a probability of less than 5\% of falling below Blim |



Figure 2. BM1. SSB from the developed SAM model and associated, derived Biomass reference points MSY $B_{\text {trigger }}$ ( 1,882 tonnes, in blue), $\mathrm{B}_{\text {ра }}\left(1,882\right.$ tonnes, dotted line) and $\mathrm{Blim}^{(1,344 \text { tonnes, dashed line). }}$


Figure 3. BM1. Fishing pressure from the developed SAM model and associated, derived F reference points $\mathrm{F}_{\mathrm{MS}}$ ( 0.147 in blue), $\mathrm{F}_{\mathrm{pa}}$ ( 0.592 , dotted line) and $\mathrm{Flim}_{\text {lim }}$ (1.275, dashed line).

For comparison purposes, running the same exercise on the BM2 SAM model of the plaice stock, provides different reference point estimates. Beginning with $\mathrm{B}_{\text {loss }}$ defined at 2061 tonnes (Appendix 2, Table 2, Figures 4 and 5).

Table 2. BM2 - Estimated reference points for plaice 7.fg.

| Reference Point | Value | Rationale |
| :--- | :--- | :--- |
| MSY Btrigger | 2885 t | Lower 5th percentile of BMSY; in tonnes |
| Fmsy | 0.479 | Stochastic simulations with segmented regression \& Beverton-Holt |
| FmsyLower | 0.263 | Median lower point estimates of Stochastic simulations |
| FmsyUpper | 0.480 | Median upper point estimates of Stochastic simulations |
| Blim | 2061 t | Lowest observed SSB |
| Bpa | 2885 t | Blim combined with the assessment error |
| Flim | 1.649 | F with 50\% probability of SSB less than Blim |
| Fpa | 0.480 | F with a probability of less than 5\% of falling below Blim |



Figure 4. BM2. SSB from the developed SAM model and associated, derived Biomass reference points MSY $B_{\text {trigger }}$ ( 2,885 tonnes, in blue), $\mathrm{B}_{\text {pa }}$ ( 2,885 tonnes, dotted line) and $\mathrm{B}_{\text {lim }}$ (2,061 tonnes, dashed line).


Figure 5. BM2. Fishing pressure from the developed SAM model and associated, derived F reference points $\mathrm{F}_{\mathrm{MSY}}$ ( 0.479 in blue), $\mathrm{F}_{\mathrm{pa}}$ ( 0.480 , dotted line) and $\mathrm{F}_{\text {lim }}$ (1.649, off scale).

## Conclusions

There is clearly a notable influence coming from the choice of natural mortality in the estimation of reference points. For the BM1 model seen by the benchmark process WKNSCS, the MSY $\mathrm{B}_{\text {trigger }}$ reference point is notably higher, while $\mathrm{B}_{\text {lim }}$ and $\mathrm{B}_{\text {trigger, }}$, lower than BM 2 . F reference points are all lower.

The premise for this is that BM1, with lower natural mortality values taken from the Irish Sea, a higher proportion of total mortality in the SAM assessment model is being attributed to fishing pressure and therefore suggest that $F$ has been relatively high and has been limiting stock development. It suggests if fishing pressure were reduced a large development of the stock would result, and a high yield even at a much lower F than is currently observed.

The larger impact on the advice is therefore the much lower estimate of $\mathrm{F}_{\mathrm{MSy}}$ in BM 1 , which comes directly from the lower natural mortality assumption and creates a significantly more pessimistic estimate of Fbased stock status than BM2.

BM2, with higher natural mortality estimates, attributes a greater proportion of total mortality to natural mortality than fishing mortality, and therefore suggest that $F$ is not limiting the stock, with $F$ below $F_{M S Y}$ since around 2007. To maximise yield it would be necessary to increase fishing pressure to the levels seen pre 2005, for a landing yield of around 1000t (similar to what has been achieved). The perception of this secondary model is much closer to the recent year's perception (Figure 5; ICES 2021b).

The selection of natural mortality values is evidently important on the calculation of reference points. The two SAM assessments, BM1 and BM2, both give potential assessments, with BM2 giving better retrospective patterns, with lower variability. This may be justification for BM2 over BM1. More work is recommended however, on the choice of natural mortalities. This should include comparison of similar plaice stocks, further review of stock-specific raw data and review of any ecosystem modelling pertaining to fish natural mortalities.


Figure 5. 2021 perception of stock status for plaice in divisions 7.f and 7.g (ICES 2021). Fishing pressure on the stock was below $F_{\text {MSy }}$ and biomass above MSY $B_{\text {trigger }}$. The short orange lines in the relative biomass plot indicate the average values of the respective years (2017 to 2019 and 2020 to 2021; ICES 2021b).

## References

ICES (2017). Report of the Workshop to consider F $_{\text {MSY }}$ ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

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# Appendix 1. Reference point calculation R-Markdown - from Base Stock assessment 

File: CS_PLE_RePt_WKNSCS_3_updated.html


CS_PLE_RePt_WKNS
CS_3_updated.htm

Appendix 2. Reference point calculation R-Markdown - from developed Stock assessment

File: CS_PLE_RePt_WKNSCS_4_updated.html


CS_PLE_RePt_WKNS
CS_4._updated.html

Response to further Review of Reference Points for Celtic Sea Plaice Divisions 7.f and 7.g
Reference points are presented for two assessments with differing natural mortalities:
(1) Ms from the Irish Sea plaice assessment (plaice7fg_001) and
(2) Ms from the English Channel plaice assessment (Ple.7fg_disc).

The first set of reference points correspond to the assessment put forward by the WKNSCS benchmark but the second (based on higher Ms ) correspond better to the previous perception of the stock (based on a SPiCT assessment). It is not clear which set of reference point are being proposed, and no reasoning is presented regarding the choice of $M$.

This document therefore reviews only the reference point calculations, although it is important to ensure the same set of Ms are used for both assessment and calculation of reference points.

Response: Review of the reference points post Benchmark revealed the issue around choice of natural mortality in estimating reference points. Both situations are presented in this WD to detail the issue and the effect of the choice. Work is ongoing to determine the choice of $M s$ for application in assessment and advice.

## Reference point calculations:

$\mathbf{B}_{\text {lim }}$ : The classification as Type 5 seems appropriate, with $B_{\text {lim }}=B_{\text {loss }}$ as specified in the guidelines.
$B_{p a}: B_{p a}$ is calculated assuming the default $\sigma=0.2$, in accordance with the guidelines. This seems appropriate given the standard deviation of $\ln (S S B)<0.2$ for both assessments.
$F_{\text {Msy }}$ : Biological parameters are derived by default from the last 10 years of data. Selectivity is derived from the last 5 years, which is justified due to trends in recent years. Assessment error parameters are taken as $F_{c v}=0.212$ and $F_{\text {phi }}=0.423$. These are stated as ICES defaults although this should be checked as default values are not given in the ICES guidelines.

Response: Fcv in the reference points presented in this updated WD are now calculated as sd(fbar)/mean(fbar). Fph=0.423 commonly used in Celtic Sea, e.g cod 7a and Haddock 7a. Comes from ICES, 2017: WKMSYREF4:
ICES. 2017. Report of the Workshop to consider FMSY ranges for stocks in ICES categories 1 and 2 in Western Waters (WKMSYREF4), 13-16 October 2015, Brest, France. ICES CM 2015/ACOM:58. 187 pp.

Stock recruitment is taken as a mixture of a segmented regression (with breakpoint estimated close to $B_{\text {lim }}$ and within the range of observed SSBs) and Beverton-Holt. $F_{M S Y}$ and the $F_{M S Y}$ range have not been checked against $F_{P .05}$ (i.e., it has not been evaluated that $P\left(S S B<B_{l i m}\right) \leq 5 \%$ when applying the ICES $A R$ ).

Response: This has now been checked, and does not change reference points.
MSY $B_{\text {trigger: }}$ For the first set of reference points (based on plaice7fg_001) the stock has not been fished at or below $\mathrm{F}_{\text {MSY. }}$ MSY $\mathrm{B}_{\text {trigger }}$ is therefore taken as $\mathrm{B}_{\mathrm{pa}}$ as specified in the guidelines. For the second set of reference points (based on Ple. $7 \mathrm{fg} g_{-}$disc) MSY $B_{\text {trigger }}$ is also taken as $B_{p a}$, although in this case $F$ has been below $F_{\text {MSY }}$ since around 2007 so the $5^{\text {th }}$ percentile of $B_{\text {MSY }}$ could be considered.

Response: This biomass is less than Bpa, so should not be used.
$F_{\text {lim: }}$ Stock recruitment was taken as a mixture of a segmented regression with breakpoint estimated and a Beverton-Holt. This is not in accordance with the guidelines which specify that a segmented regression with the point of inflection at $B_{\text {lim }}$ should be used (although the BP is estimated close to $B_{\text {lim }}$ ).

Response: Now corrected to use Hockeystick with breakpoint at Blim for the estimation of Flim. Slightly increases Flim and Fpa

Assessment and advice error were (correctly) not included, and all other settings were the same as for $\mathrm{F}_{\text {MSY }}$.
$F_{p a}: F_{p a}$ was derived as $\sim F_{\text {lim }} / 1.4$ using the $F p a()$ function in the icesAdvice package. This appears to be a legacy function from the previous guidelines as Table 16.4.3.1.1 now specifies that $F_{p a}$ should be taken as $F_{\text {P. } 05}$ (i.e., the $F$ that would lead to $P\left(S S B \geq B_{\text {lim }}\right)=95 \%$ in the ICES $A R$ ).

## Annex 4: Audits

## Audit of Cod (Gadus morhus) in Division 6.a (West of Scotland)

Reviewer: Dave Stokes 4/6/2022

## General

For single-stock summary sheet advice:

1. Assessment type: Benchmarked in 2020. The current assessment is in line with the stock annex.
2. Assessment: Analytical age-based SAM assessment.
3. Forecast: Short-term stochastic forecast carried out in SAM.
4. Assessment model: SAM assessment with catch data tuned with five survey series.
5. Data Issues: Q1 survey data not available for interim recruitment; low sampling from OTB-CRU resulting in biased estimates for ages $1-2$ so removed from assessment for 2021; VMS data unavailable so misreporting estimated from Marine Scotland Compliance data.
6. Consistency: Other than data issues above, recruitment assumption reduced to five years rather than ten years, otherwise assessment is consistent.
7. Stock status: SSB is estimated to be below all reference points and F estimated to be above.
8. Management plan: There is no agreed shared management plan with UK for this stock, and ICES provides advice according to ICES MSY approach - zero catch advice.

## General comments

The report is comprehensive and easy to follow with only a few minor comments. The assessment itself has some updates driven by missing data as well as some noise in tuning indices for older ages and recent years in particular. Overall however, the assessment is consistent with the stock annex and guidelines for retrospective bias and therefore accepted by WGCSE 2022.

## Technical comments

Inclusion of survey variance in the assessment is interesting and uncommon so some description in the stock annex on how it was introduced and how it is used, etc. would be useful as well as listing in the inputs to SAM in SA/Advice sheet.

Check minor comments Tables $6,7 \& 9$ in advice sheet.

## Conclusions

The assessment is well presented, supported by the data and the report, and appears to have been carried out either as prescribed or as documented by the stock coordinator in the report and advice sheet.

## Audit of Cod (Gadus morhua) in Division 7.a (Irish Sea)

Reviewer: Sara-Jane Moore

## General

For single-stock summary sheet advice:

1. Assessment type: stock was benchmarked in February 2022. Assessment run based on the benchmark but without recreational removals.
2. Assessment: age-based.
3. Forecast: yes, as short-term stochastic projections.
4. Assessment model: SS3.
5. Consistency: last year's assessment was based on Survey trends using the biomass index of the NIGFS-WIBTS-Q1 survey. Last year stock was Category 3 .
6. Stock status: Fishing pressure on the stock is below FmSy and spawning-stock size is below MSY Btrigger, $\mathrm{B}_{\mathrm{pa}}$, and $\mathrm{Blim}_{\text {lim }}$
7. Management plan: No agreed management plan.

## General comments

The assessment was benchmarked in 2022 and changed from category 3 to category 1 (ICES, 2022b), therefore historical assessments are not comparable. Additional information included an updated natural mortality M , inclusion of multiple survey indices, commercial catch and catch-at-age information and recreational removals. Owing to uncertainty around calculations of the recreational removals and selectivity pattern, all recreational removals have been excluded in the 2022 assessment. This makes minor changes to the perception of the stock.

## Technical comments

- Catch SOP check for 2020 is low (0.74525). when you multiply lnaa by lwaa get 265581 tonnes not 181 tonnes as in landings Lowestoft file.
- Get Mohns rho of -0.14 for F.
- $\quad$ Fleet 2 (NIGSQ1) has an error of 0.96 in 2002, this seems quite large.

What was the basis of the S-R relationship? It does not look appropriate from SS3 plots.

## Conclusions

Depending on input values corrections to be checked, the stock assessment appears to have been done correctly. If current Lowestoft files are used for forecast, then the forecast will have to be re-run. The perception of the stock is unlikely to change however.

Minor edits suggested for report.

## Audit of Cod (Gadus morhua) in divisions 7.e-k (eastern English Channel and southern Celtic Seas)

Date: 20/05/22
Auditor: Paul Bouch

## General

ICES advises for zero catch in 2023. Fishing pressure is above $F_{\text {MSY }}$ and $F_{p a}$ and spawning-stock size is below all relevant reference points. The TAC for 2021 was significantly overshot with discards estimated to be $54 \%$.

For single stock summary sheet advice:

1. Assessment type: update.
2. Assessment: analytical.
3. Forecast: presented.
4. Assessment model: SAM - tuning using a VAST index (combining EVHOE and IRGFS) and the FR-OTDEF commercial index.
5. Data issues: No issues.
6. Consistency: Similar to last year's assessment. Mohn's Rho values are high but similar in magnitude to previous years.
7. Stock status: SSB continues to be below $\mathrm{B}_{\mathrm{pa}}, \mathrm{Blim}_{\text {lim }}$ and $\mathrm{B}_{\text {trigger. }}$. Fishing pressure declined slightly to drop below Flim since last year, but it remains well above $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{F}_{\mathrm{ms}}$.
8. Management Plan: ICES recommends zero catch advice. Mixed fisheries advice will be provided later in the year.

## General comments

The assessment is available on stockassessment.org and is in agreement with what is specified in the stock annex.

## Technical comments

No issues were identified.

## Conclusions

The assessment has been performed correctly. The stock assessment for this stock has been audited internally within WGCSE, and no issues were found.

# Audit of Haddock (Melanogrammus aeglefinus) in Division 7.a (Irish Sea) 

Date: 31/06/2022
Auditor: Andrzej Jaworski

## General

ICES provides annual catch advice for this stock based on the MSY approach. A full analytical assessment and forecast were performed in 2022 in accordance with the procedures outlined in the stock annex. The assessment is based on an age-structured model.

For single stock summary sheet advice:

1. Assessment type: Update. Age-structured assessment. The stock was benchmarked by WKIrish in 2017.
2. Assessment: Age-structure assessment. Stock Category 1.
3. Forecast: Short-term forecast is presented. Conducted using FLR libraries. The introduction of ASAP has considerably changed the catch advice compared to previous years.
4. Assessment model: Age-structured assessment model using Age-Structured Assessment Program (ASAP) with commercial catches and four survey indices.
5. Consistency: There is close agreement of the stock trends in the current assessment and the benchmark assessment.
6. Stock status: Spawning-stock biomass (SSB) is at relatively high levels in the time-series and above MSY $\mathrm{B}_{\text {trigger. }}$ but it has declined in recent years. Fishing mortality ( F ) has been below Fmsy since 2012. The stock is characterized by highly variable recruitment. Recent recruitment has been low (and below the time-series mean).
7. Man. Plan: There is no specific management plan for the stock.

## General comments

The assessment was conducted correctly and it followed the methods detailed in the stock annex.

## Technical comments

ASAP analysis was correctly performed.

## Conclusions

The assessment has been performed correctly and provides an appropriate basis for providing catch advice.

## Checklist for review process

## General aspects

- Has the EG answered those ToRs relevant to providing advice? Yes
- Is the assessment according to the stock annex description? Yes
- Is general ecosystem information provided and is it used in the individual stock sections. Yes
- If a management plan has been agreed, has the plan been evaluated? No


## For update assessments

- Have the data been used as specified in the stock annex? Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex? Yes
- Is there any major reason to deviate from the standard procedure for this stock? No
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice? Yes


# Audit of Haddock (Melanogrammus aeglefinus) in divisions 7.b-k (southern Celtic Seas and English Channel) 

Date: 03/06/2020
Auditor: Ruth Kelly

## General

This stock was benchmarked by WKCELTIC in 2020, and assessment methods have been updated according to guidance from this benchmark assessment as detailed in the stock annex.

For single stock summary sheet advice:

1. Assessment type: Update.
2. Assessment: Analytical.
3. Forecast: Presented.
4. Assessment model: State-Space Assessment Model (SAM).
5. Data issues: Due to COVID-19 catch sampling of this stock was disrupted in 2020. Catch sampling in 2021 increased compared to 2020 and is considered sufficient to describe the stock.
6. Consistency: This assessment follows the assessment method (SAM) that is described in the 2020 WKCELTIC benchmark of this stock, and has been applied since 2021.
7. Stock status: Fishing pressure on the stock is below Fmsy and spawning-stock size is above MSY $B_{\text {trigger, }} \mathrm{B}_{\mathrm{pa}}$, and $\mathrm{Blim}_{\text {l }}$.
8. Management Plan: EU multiannual management plan (MAP) for the Western Waters (EU, 2019).

## General comments

Assessment is well documented in advice sheet, report and stock annex.

## Technical comments

Data have been provided via ICES SharePoint, and SAM assessment based on VAST modelled time-series are available on stockassessment.org.

## Conclusions

Report, advice sheets and analysis are available, and analysis appears to have been performed correctly.

## Checklist for audit process

## General aspects

- Has the EG answered those ToRs relevant to providing advice? Yes
- Is the assessment according to the stock annex description? Yes
- If a management plan is used as the basis of the advice, has this been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary? Yes
- Have the data been used as specified in the stock annex? Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex? Yes
- Is there any major reason to deviate from the standard procedure for this stock? No
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice? Yes


# Audit of Megrim (Lepidorhombus spp.) in divisions 4.a and 6.a (northern North Sea, West of Scotland) 

Reviewer: Simon Fischer, 24/05/2022

## General

The assessment and forecast for megrim in 4.a and 6.a are complex but reproducible.
Short description of the assessment:

1. Assessment type: Update.
2. Assessment: Accepted.
3. Forecast: Accepted.
4. Assessment model: Custom Bayesian surplus production (Schaefer) model with six survey indices, some of them from delta-gamma GLMs. Discards before 2013 are estimated but used as input for the assessment model.
5. Consistency: High consistency of estimates relative to MSY reference points. Absolute values (used for forecast) more variable.
6. Stock status: $\mathrm{B}>\mathrm{MSY} \mathrm{B}_{\text {trigger }}$ and $\mathrm{F}<\mathrm{F}_{\text {msY. }}$
7. Management plan: EU multiannual plan (MAP) for Western Waters and adjacent waters, using Fmsy ranges and MSY Btrigger.

## General comments

The report is concise and could benefit from more detailed descriptions of the assessment and forecast procedure and outcomes.

The Stock Annex (SA) would benefit from a rewrite and update. There is a lack of model specifications for the assessment model, and the description for the short-term forecast is extremely short and does not give details about how to perform it.

The stock assessment method (a Schaefer surplus production model) is used as a category 1 datarich stock assessment model since 2012. Since then, the ICES system has changed and such an assessment would likely be considered as a category 2 or 3 data-limited approach nowadays. It is unknown if this assessment configuration would meet the strict acceptance criteria for a category 2 surplus production model.

## Technical comments

The assessment is run with a Bayesian model in WinBUGS, which is very slow, particularly for a surplus production model. Also, the forecast takes $>0.5$ hour to run. The assessment is conducted with numerous R/Markdown scripts and appears overly complex with several versions run concurrently, which makes auditing this stock challenging. The assessment would benefit from cleaning the assessment procedure, e.g. by creating an R package for the model, which would also remove the possibility of accidentally introducing mistakes.
The assessment could be reproduced.
The forecast is conducted by running a wide range of discrete catch options and then selecting them based on catch, F, biomass, or risk considerations. The forecast is stochastic with many thousand iterations, but metrics such as biomass or F are reported as the mean (not median) of the distributions. There is no documentation (SA or report) on how to derive the scenarios shown in the advice sheet from the numerous catch scenarios. The derivation of the final forecast is somewhat intransparent but follows the same procedure as in previous years.

## Conclusions

The assessment is complex but consistent with the approach used in previous years. Full compliance with the stock annex cannot be evaluated because the stock annex is vague and does not give comprehensive details about the data, assessment and forecast procedures.

## Audit of Plaice (Pleuronectes platessa) in Division 7.a (Irish Sea)

Reviewer: Claire Moore, 01/06/2022

## General

The assessment and forecast were performed correctly.
Short description of the assessment:

1. Assessment type: Update.
2. Assessment: Accepted.
3. Forecast: Accepted.
4. Assessment model: SAM assessment: landings + discards + three survey indices.
5. Consistency: Same approach as in previous years; downscaling in SSB in retro.
6. Stock status: Fishing pressure below FMSY and spawning-stock size is above MSY Btrigger, $\mathrm{B}_{\mathrm{pa}}$, and $\mathrm{B}_{\mathrm{lim}}$.
7. Management plan: Bycatch stock within EU MAP, no agreed management plan with UK.

## General comments

None.

## Technical comments

Intersessional work should look into the fact that an increasing proportion of the catch is contained in the plus group, which may be increasing the uncertainty in the SSB of the model. $47 \%$ of catch in last five years has been contained in $8+$ group.

Future work should also consider the increasing evidence of substock structure within the population, which would include growth rates and mortality considered within the model.

## Conclusions

The assessment and forecast for plaice in 7. a could be reproduced and were performed correctly. The report was well written, and concise, with very little feedback to address.

# Audit of Plaice (Pleuronectes platessa) in Division 7.e (western English Channel) 

Reviewer: Vladimir Laptikhovsky

## General

The 7.e plaice is relatively data-rich for an ICES category 3 data-limited stock. There are no major concerns for this stock to provide advice following category 3 methods. However, the stock might be upgraded in the future.

For single-stock summary sheet advice.
Short description of the assessment as follows:

1. Assessment type: update of the annual assessment.
2. Assessment: accepted.
3. Forecast: not presented.
4. Assessment model: choice of rfb rule was based on the new ICES technical guidelines for stocks in categories 2 and 3 and supported by expert group. This harvest control rule accounts for life-history parameters and the previously advised catch. Legacy XSA assessment and exploratory SAM assessment were carried out but not considered.
5. Consistency: last year assessment used the 2 over 3 rule and was based on the SSB estimate of a landings' only XSA assessment. This assessment is not used anymore for the ICES advice because the rfb rule is applied to empirical data.
6. Stock status: Lmean < L mSY proxy, Biomass index > MSY B trigger. $^{\text {t }}$
7. Management plan: ICES is aware of the multiannual management plan which has been adopted by the EU for this stock and which ICES considers to be precautionary. There is no agreed shared management plan with UK for this stock, and ICES provides advice according to ICES MSY approach.

## General comments

The stock unit (Division 7.e) does not correspond with the management unit (divisions 7.d and 7.e), and this divisional mismatch hampers the effective management of plaice in the Western English Channel. However, some provision must be made to consider the effective management of adjacent plaice stocks, given that components of the $7 . e$ stock are also taken during the spawning period in 7.d.

## Technical comments

A degree of uncertainty exists over the landings' statistics for this stock, given that mature plaice migrate between 7.d and 7.e during the spawning period. The assessment results depend on the mixing rate assumption estimated from historical tagging data that apply a spawning migration correction that reallocates $15 \%$ of Q1 landings for the mature proportion of the catch from 7.d to 7.e.

## Conclusions

The 2021 stock assessment for this stock has been audited internally within WGCSE, and no issues were found.

## Audit of Plaice (Pleuronectes platessa) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea)

Reviewer: Claire Moore, 30/05/2022

## General

The assessment was performed correctly, there are some checks required on the forecast.
Short description of the assessment:

1. Assessment type: Cat3 rfb rule.
2. Assessment: Accepted.
3. Forecast: Not applicable.
4. Assessment model: RFB rule - using a survey biomass index for trends.
5. Consistency: Benchmark model (ICES 2022) rejected due to error in some model assumptions.
6. Stock status: The stock is above Fmsy proxy and below MSY Btrigger.
7. Management plan: EU MAP, but no agreed management plan with the UK.

## General comments

A number of updates have been requested to the report (see comments) and the stock annex before publication.

## Technical comments

The rfb rule has been applied correctly, the advice sheet follows the code provided.

## Conclusions

The assessment for ple in 7.fg could be reproduced and were performed correctly.

# Audit of Plaice (Pleuronectes platessa) in divisions 7.h-k (Celtic Sea South, southwest of Ireland) 

Reviewer: Vladimir Laptikhovsky

## General

The 7.h-k plaice is an ICES category 3 data-limited stock. There are no major concerns for this stock to provide advice following category 3 methods.

For single-stock summary sheet advice
Short description of the assessment as follows:

1. Assessment type: update of the annual assessment.
2. Assessment: accepted.
3. Forecast: not presented.
4. Assessment model: choice of rfb rule was based on the new ICES technical guidelines for stocks in categories 2 and 3 and supported by expert group. A survey biomass index was produced using VAST biomass index of the VAST model and applied as an indicator of stock development. The advice is based on the recent advised catches, multiplied by the ratio of the mean of the last two index values (index A) and the mean of the three preceding values (index B), a ratio of observed mean length in the catch relative to the target mean length, a biomass safeguard, and a precautionary multiplier. The stability clause was considered and applied since the new catch advice was more than $20 \%$ higher than the recent advised catches.
5. Consistency: last year assessment is based on the rfb rule to provide MSY advice. A survey combined biomass index was used as an indicator of stock development. The advice is based on the ratio of the mean of the last two index values (index A) and the mean of the three preceding values (index B), multiplied by the recent catches, a ratio of observed mean length in the catch relative to the target mean length, a biomass safeguard, and a precautionary multiplier. The stability clause was considered and applied.
6. Stock status: Lmean $>\mathrm{L}$ MSY proxy, Biomass index $>$ MSY $B_{\text {trigger. }}$
7. Management plan: The EU multiannual plan (MAP) for stocks in the Western Waters and adjacent waters applies to bycatches of this stock. There is no agreed shared management plan with UK for this stock, and ICES provides advice according to ICES MSY approach.

## General comments

The stock unit (divisions 7.h-k is consistent with the assessment area. However, to date no stock identification studies (including tagging) have been conducted on plaice in $7 \mathrm{~h}-\mathrm{k}$, which is on the southwestern margins of the species distribution. There is evidence in other areas to suggest that plaice is a highly mobile species, and therefore it is possible that ple.27.7.h-k might be an extension of larger adjoining populations, but tagging and genetic would need to be completed to determine this.

## Technical comments

The discard rates for this stock are highly variable over time being mean $42 \%$ in 2004-2019. This stock should be considered for the next SPiCT workshop to assess if it can be moved to a category 2 stock.

## Conclusions

The 2022 stock assessment for this stock has been audited internally within WGCSE, and no issues were found.

# Audit of Pollack (Pollachius pollachius) in subareas 6-7 (Celtic Seas and the English Channel) 

Reviewer: Pia Schuchert, 31/05/2022

## General

For single stock summary sheet advice:

1. Assessment type: update from last year assessment.
2. Assessment: depletion corrected average catch.
3. Forecast: No.
4. Assessment model: DCAC from NOAA toolbox for category 4 stocks.
5. Data issues: The available data consist purely of commercial landings, which is also reflected in the fact that the stock is in a data-limited category. In previous years, exploitation has been consistently below the DCAC 95\% confidence interval in both management areas.
The amount and impact of a recreational fishery is currently not included, but efforts are made to include this in the future.
6. Consistency: Same approach as last year.
7. Stock status: The available information on the stock is insufficient to evaluate the exploitation rate and stock trends for Pollack.
8. Management Plan: The European Parliament and the Council have published a multiannual management plan (MAP) for the Western Waters (EU, 2019). This plan applies to demersal stocks including pollack in ICES subareas 6 and 7. ICES advises on the precautionary basis. There is no agreed shared management plan with UK for this stock.

## General comments

The report contained the expected information in a clear manner and is very thorough.
The advice is on precautionary basis as under the ICES framework for data-limited stocks.

## Technical comments

None.

## Conclusions

The assessment has been performed correctly.

## Audit of Sole (Solea solea) in Division 7.a (Irish Sea)

Reviewer: Simon Fischer, 24/05/2022

## General

The assessment and forecast were performed correctly.
Short description of the assessment:

1. Assessment type: Update.
2. Assessment: Accepted.
3. Forecast: Accepted.
4. Assessment model: XSA (in FLR) - tuning by one survey index.
5. Consistency: Same approach as in previous years; some rescaling is appearing.
6. Stock status: SSB has been increasing and is just above MSY $B_{\text {trigger }}$ in 2022; F has also been increasing and is now above $\mathrm{F}_{\text {msy }}$ and $\mathrm{F}_{\mathrm{pa}}$.
7. Management plan: None.

## General comments

None.

## Technical comments

Discards are not used in the assessment (but considered in the advice) and have been increasing recently ( $\sim 12 \%$ in 2021). Future benchmarks should consider the inclusion of discards into the assessment, particularly because of the recent changes in the discarding behaviour.

Some of the characteristics seen in this year's assessment (rescaling of SSB in recent years, estimated number of age 8+ fish in the stock in 2020 being higher than the combined numbers for ages 7 and $8+$ in 2019) are likely a feature of XSA and its formulation. Using an alternative stock assessment model, such as the state-space model SAM, might help to avoid this behaviour and should be considered in a future benchmark.

## Conclusions

The assessment and forecast for sole in 7.a could be reproduced and were performed correctly.

Audit of Sole (Solea solea) in divisions 7.b and 7.c (West of Ireland)

Section currently missing

# Audit of Sole (Solea solea) in Division 7.e (western English Channel) 

Reviewer: Sofie Nimmegeers, 30/05/2022

## General

The assessment and forecast were performed correctly.
Short description of the assessment:

1. Assessment type: Update.
2. Assessment: Accepted.
3. Forecast: Accepted.
4. Assessment model: XSA (in FLR) - tuning by four tuning fleets (two fishery-independent surveys: UK-FSP and Q1SWBeam; and two commercial lpue time-series: UK-CBT-late and UK-COT).
5. Consistency: Same approach as in previous years; The retrospective has been increasing slightly in recent years; however, this does not cause concerns because the SSB is usually underestimated while the fishing mortality is overestimated.
6. Stock status: SSB has been increasing and is well above MSY $B_{\text {trigger, }}, \mathrm{B}_{\mathrm{pa}}$ and $\mathrm{B}_{\text {lim }}$ (highest value of the time-series in 2022); F has also been increasing and is now at $\mathrm{Fmsy}^{\text {ms. }}$
7. Management plan: EU MAP, but not accepted by all clients

## General comments

The report is well-structured and contains a series of very informative graphs.
A few small comments:
*please adjust the last column of the table with the XSA assessment settings used at the last three working groups. It should be referring to 'WGCSE 2022' instead of 'WGCSE 2021'
*In Section 32.9: Uncertainties in assessment and forecast - discards, it is stated that 'Discarding is considered negligible in the sole fishery, averaging only $0.44 \%$ of the total international catch weight in 2021.' The $0.44 \%$ is the average discard rate of 2019-2021 and not of 2021.

## Technical comments

Discards are not used in the assessment (but considered in the advice). As discards are low (average discard rate of $2019-2021=0.44 \%$ ), due to the implementation of the landing obligation in 2016, it is unlikely to become a problem in the future. Nevertheless, as the time-series on discards increases, a future benchmark might look into estimating historical discards.

The current XSA assessment has a "taper range" of 15 years, which means all survey data older than 15 years are ignored and data from 7-15 years ago are strongly down-weighted. The scientific survey index time-series are increasing in length and early years are being cut off due to the taper range. Therefore a re-evaluation of assessment parametrisation should be considered in a future benchmark. The inclusion of commercial tuning indices should also be questioned.

## Conclusions

The assessment and forecast for sole in $7 . e$ could be reproduced and were performed correctly.

# Audit of Sole (Solea solea) in divisions 7.f and 7.g (Bristol Channel, Celtic Sea) 

Reviewer: Timothy Earl, 27/05/2022

## General

The assessment was performed correctly, there are some checks required on the forecast.
Short description of the assessment:

1. Assessment type: Update.
2. Assessment: Accepted.
3. Forecast: Accepted.
4. Assessment model: SAM using landings and discards, tuned using one survey at-age and five biomass commercial LPUE series.
5. Consistency: Since the benchmark in 2020, the current assessment has shown good consistency.
6. Stock status: The stock is above MSY $B_{\text {trigger. }}$ Fishing mortality is at $\mathrm{F}_{\text {mSY. }}$
7. Management plan: EU MAP, but not accepted by all clients.

## General comments

None.

## Technical comments

The stock is described (correctly) as being at Fmsy because the numbers are the same to two significant figures. If this departure from the standard ICES description is not acceptable, it is also correct to say $\mathrm{F}_{2021}(0.2504)$ is below $\mathrm{FMSY}(0.251)$.

Some of the other advice sheets are presenting the MAP range as just the lower/upper point, so " $0.136-0.251$ " would be written as "" 0.136 ".

Some of the inputs to the short-term forecast don't seem to match between the advice sheet and the file "stf_input.csv". From this file the sum of products of the N, Mat and stock weight give SSBs of 6095 in 2022, and 5896 in 2023. The corresponding number in the advice sheet are 6221 in 2022 (Table 9) and 6120 (Table 1).

## Conclusions

The assessment for sole in 7.fg could be reproduced and were performed correctly. There are minor issues in the forecast to review.

# Audit of Sole (Solea solea) in divisions 7.h-k (Celtic Sea South, southwest of Ireland) 

Reviewer: Timothy Earl, 27/05/2022

## General

The assessment was performed correctly.
Short description of the assessment:

1. Assessment type: No assessment, category 5 advice.
2. Assessment: N/A.
3. Forecast: N/A.
4. Assessment model: N/A.
5. Consistency: Same approach as in last year. No change to advice.
6. Stock status: Unknown.
7. Management plan: EU MAP applies, but not relevant to category 5.

## General comments

None.

## Technical comments

Due to poor knowledge of the stock identity, poor sampling from Division 7.h and potential issues with historic misreporting, this stock is categorised as category 5 (catch data only). Discards are low (less than $1 \%$ ).

The report section on the Landings Obligation is written referring to its future implementation, it may be worth reviewing this paragraph to reflect how it has affected the fishery.

By providing advice on based only on the catch trends, it seems to me that we are ignoring potentially important data on the abundance of sole in this area that could be used to provide more robust advice. I suggest that the stock assessor consider applying the VAST survey methodology in a similar way to the assessment of plaice in this area for consideration at future Working Groups, along with reviewing whether recent length data sampling is sufficient for the `rfb` approach to be applied.

## Conclusions

The advice for sole in $7 . \mathrm{h}-\mathrm{k}$ could be reproduced and was performed correctly.

# Audit of Whiting (Merlangius merlangus) in Division 6.a (West of Scotland) 

Reviewer: Emilie Le Luherne, 25/05/2022

## General

For single stock summary sheet advice:

1. Assessment type: update from last year assessment.
2. Assessment: analytical age-based.
3. Forecast: Presented, in form of short-term stochastic projections.
4. Assessment model: SAM with catch and survey data, SURBA assessment for comparison.
5. Data issues: Total landings (nominal landings, ICES statistics) in 2021 were the highest in the last 18 years and constituted $91 \%$ of the TAC for that year. The sampling levels in 2021 in the Scottish fleet (taking the majority of the catch), similarly to 2020, were lower compared to previous years due to COVID-19. This WGs estimates of discards have been based on data collected in the Irish and Scottish discard programme and raised by landings. Mean weights-at-age 1 in discards were considerable higher compared to previous years, but this was most likely an effect of the low number of samples from the Nephrops fleet.
6. Consistency: Same approach as last year. The 2022 assessment thus runs up to 2021, deviating from the model configuration agreed in the benchmark. The 2021 assessment model was revised in 2022 due to an error in stock mean weights-at-age data inputs which affected SSB estimates, but not the published catch advice (the revised estimates for last year's assessment are as follows: F (1-3) in $2020=0.065$, SSB $_{2021}=30357 \mathrm{t}$ ). The residuals were not substantially affected by the updates made to the model, showing similar patterns to last year's assessment model. The retrospective patterns show a downward revision of SSB and recruitment, and upward revision of F. Both SSB and F remain in their respective envelopes and SSB is above MSY $\mathrm{B}_{\text {trigger }}$ and F consistently below Fmsy.
7. Stock status: SSB is considered to be close to the 1990s state and above MSY B trigger, increasing in recent years. Fishing mortality has been consistently below FmsY since 2005.
8. Management Plan: There is no agreed shared management plan with UK for this stock, and ICES provides advice according to ICES MSY approach. The EU map takes discards of this stock into account.

## General comments

The report contained the expected information in a clear manner and is very thorough. A complementary assessment in SURBAR is provided, which supports the findings from the SAM assessment.

The advice is on MSY basis for the second time after years of precautionary advice.

## Technical comments

None.

## Conclusions

The assessment has been performed correctly.

# Audit of Whiting (Merlangius merlangus) in divisions 7.b-c and 7.e-k (southern Celtic Seas and eastern English Channel) 

Reviewer: Helen Dobby 27/05/2022

## General

For single stock summary sheet advice:

1. Assessment type: There was an IBP in 2021. This year's assessment follows the approach agreed at the IBP (although stock annex not updated following IBP, so difficult to verify this).
2. Assessment: analytical age-based.
3. Forecast: Presented, in form of short-term stochastic projections.
4. Assessment model: SAM with catch and survey data.
5. Data issues: No major data issues.
6. Consistency: Assessment follows approach agreed at IBP although some discussion of removal of LPUE which results in improved diagnostics (retro bias/Mohn's rho), but considered too significant to change out with a benchmark. Forecast assumptions of recruitment modified to shorter resampling period (2015 onwards, rather than 2010) due to apparent reduction in recruitment in recent years.
7. Stock status: SSB is considered below reference points and F above reference points. SSB in forecast very sensitive to assumption regarding recruitment. Short-term resampled recruitment implies stock is below $B_{\lim }$ in forecast and therefore zero catch advice.
8. Management Plan: There is no agreed shared management plan with UK for this stock, and ICES provides advice according to ICES MSY approach - zero catch advice.

## General comments

There was no up to date Stock Annex available for this stock so it was quite difficult to ascertain whether the assessment had been conducted correctly (as agreed). The audit therefore consisted of cross-checking report tables with data in SA.org, re-running the assessment and cross-checking output from SA.org with the WG report and advice sheet. ('RevRec' stock assessment on SA.org and whg.27.7b-ce-k_DS.doc advice sheet were checked).
The report was concise, well-written and easy to follow. It would be useful to present additional data in the report - such as mean weights-at-age in landings and discards, and also the landings fraction (or landings and discard numbers-at-age) since these quantities are used in SA.org in the forecast.

The assessment is borderline acceptable in terms of retro bias/Mohn's rho with significant revisions in SSB and F occurring with the addition of extra years of data.

## Technical comments

There were a number of discrepancies between the WG report and the assessment data in SA.org which need to be checked by the stock assessor:

- Table 6: stock weights (see WG report for details) mismatch in 2019, 2020 and 2021 missing from WG report.
- Natural mortalities also differ between WG report and SA.org.

In addition, some numbers in the forecast need to be corrected (in both the WG report and advice sheet) and/or checked:

- The intermediate year/forecast assumptions table needs to be updated to account for the shortened recruitment time-series assumption
- The $\mathrm{F}=\mathrm{F}_{2022}$ and $\mathrm{SSB}_{2024}=$ SSB $_{2023}$ options in SA.org appear to have different intermediate year $F$ assumptions to the other options (perhaps TAC constraint and/or $\mathrm{F}_{\mathrm{sq}}$ rather than $F_{\text {mix }}$ ).


## Conclusions

The stock assessment appears to have been done correctly (at least the WG report/advice sheet is consistent with SA.org). However, some of the assessment input data require checking (as above) and in addition some of the forecast options in SA.org appear to be incorrect.

## Annex 5: Technical Minutes of the Stony Brook University Review Group for the Advice Drafting Group for Biology and Assessment of the Celtic Sea Fisheries Resources

The following review was done by the Review Group of Data-limited Stocks (RGDLS) on June 3, 2022.

# Technical Minutes of the Stony Brook University Review Group for the Advice Drafting Group for Biology and Assessment of the Celtic Sea Fisheries Resources 

May 17 - June 6, 2022
Stony Brook University, Stony Brook, New York, USA

## Reviewers:

Dr. Cameron Hodgdon (Co-Chair), Dr. Ming Sun (Co-Chair), Dr. Hsiao-Yun Chang, Nathan Willse, Noah Khalsa , Jaeheon Kim (UMaine), Emily Fitting (UMaine), Patricia Woodruff, Katrina Rokosz, Qingqiang Ren, Tatum Eigenberger, Dr. Yunzhou Li, Xiangyan Yang, Elliot Sivel, Robyn Linner, Xindong Pan, Farrah Leone, Claire Ober, Stephanie Arsenault, \& Allegra Ervin

## Faculty Advisor:

Dr. Yong Chen (Professor, School of Marine and Atmospheric Sciences, Stony Brook University)

## ICES Secretariat:

Anne Cooper

## Review Process



The Stony Brook University Review Group (SBU RG) met on May 17th, 2022 to examine the review materials, discuss the review process, and assign individuals to a subgroup of 4 reviewers focusing on particular stocks. The relevant materials were distributed to each RG subgroup on the ICES SharePoint website. Reviews were carried out after the working group (WG) completed the final report for the WGCSE stocks. In general, the ICES guidelines for review groups (RG) were followed. The RG focused on the consistency between the WG report and the stock annex, i.e., checking whether the assessment, calculation of biological reference points (BRPs), and forecast were carried out in accordance with the stock annex, Terms of reference (ToRs), and RG guidelines. Furthermore, the RG examined the data quantity and quality, assessment method, assumptions, technical measures, uncertainty, presentations, and BRPs for each reviewed stock to ensure that management measures are based upon the best scientific information available. The RG finalized their reports on June 1st to determine the status of each group's report as well as their final decision of accepting or rejecting the assessment, and discussed any remaining issues. Table 1 lists the stocks reviewed by the SBU RG along with the suggestion (accept, accept with caveats, or reject).

Table 1. List of stocks reviewed by the SBU RG.

| Stock code/ <br> draft advice link | Assessment <br> Type | Data <br> category | Expert <br> group | RG <br> suggestion |
| :--- | :--- | :--- | :--- | :--- |
| ple.27.7e | WKLIFE X approach <br> $r f b$ rule | 3 | WGCSE | Accept |
| ple.27.7fg | WKLIFE X approach <br> $r f b$ rule | 3 | WGCSE | Accept |
| ple.27.7hk | WKLIFE X approach <br> $r f b$ rule | 3 | WGCSE | Accept with Caveats |

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## General Comments

The RG identified some common issues regarding the reports, data, and methods among the stocks, and they are summarized as follows:

The RG understands that the recent WKLIFE document outlines new methodology in the form of a flowchart that WGCSE must follow. In most cases, this was not appropriately outlined in the individual assessment reports. Discussion of this in the context of deviations from the respective stock annexes is necessary for all assessment reports.

This year is the first time that WKLIFE was applied to the data-limited bycatch species. Compared to the assessment models used in previous assessment, WKLIFE approaches have a more explicitly defined roadmap directing different rules based on available data and life history traits. However, the RG believes that performance of WKLIFE approaches are hard to evaluate without quantitative diagnostics such as retrospective analysis or relative errors. As uncertainty due to climate change, lack of knowledge in stock structure, and variations in spatio-temporal distribution are detected for the assessed stocks, it is currently unclear how robust the WKLIFE approaches are. The RG suggests further considerations of uncertainty should be introduced to the WKLIFE approaches to ensure better management effects. Overall, the RG believes the WKLIFE flowchart for assessment is too rigid and does not allow individual stock flexibility.

The RG recommends exploratory analyses of various data-limited-methods concerning WGCSE stocks currently with empirical assessments. The RG has included a list of data-limited-methods resources in the References section.

## Stock Specific Issues

The RG suggested the following stocks to be accepted as long as certain suggestions are considered ("Accept with Caveats"). These stocks and suggestions are:
ple.27.7hk

- Discussion of any exploratory SPiCT use is discussed or the lack of use is further justified.
- Because of the lack of an annex, discussion of previous management and assessment practices is necessary for the report.
- The additional general and technical comments are addressed.


## Plaice (Pleuronectes platessa) in Division 7.e (western English Channel) (ple.27.7e)

1. Assessment Type:

- Update
- Last benchmark: 2015

2. Assessment:

- Accept

3. Forecast:

- None.

4. Assessment Model:

- WKLIFE X: $r f b$ rule is used for advice
- SPiCT, XSA, and SAM assessments are explored, but results are not used for advice.


## 5. Consistency:

- The natural mortality and the maturity ogives were the same as previous assessments in Division 7.a and 7.fg. The natural mortality is from Division 7.a, which has been changed recently, but the origin values are still used in Division 7.e.
- The landing-only XSA has been updated but not used for advice anymore.
- The SAM assessment is still used this year for exploratory purposes only without a benchmark.
- The stock in Division 7.e had a joint TAC with the plaice stock in 7.d in previous year's management until 2021. This year the stock in 7.e has an individual TAC.


## 6. Stock Status:

- Length-based indicators show the stock was overfished ( $\mathrm{L}_{\text {Mean }}<\mathrm{L}_{\mathrm{F}=\mathrm{M}}$ ) during 2014-2021.
- The SAM assessment indicates strong overfishing ( $\mathrm{F}>\mathrm{F}_{\mathrm{MSY}}$ ), but it's not used for advice purposes.
- The biomass from the FSP survey has been decreasing since 2014 and had the first increase in 2021 since then.
- The total international landings of Division 7.e were 1373 t in 2020 and 1403 t in 2021 after the migration correction.
- The landings have decreased since 2017.


## 7. Management Plan:

- There are no specific management objectives except for the EU multiannual plan for the region.
- The recommended TAC for years 2023 and 2024 is 1219 tonnes each year.
- Some technical measures are applicable in 2021 and 2022, including a minimum 80 mm mesh size and a minimum 27 cm landing size.


## 8. General Comments:

- The WG report is well written and follows the stock annex. The data and data issues are well documented and discussed at length.
- The WG has thoroughly considered the issues in the connectivity between stocks 7.d and 7.e. Landings were corrected based on the understanding of migration which makes the results of the assessment acceptable.
- The WG has successfully applied the rfb rule in the assessment following the WKLIFE approach and reliable catch advice is given. Discards were included in the SAM assessment and an even worse overall overfishing situation was identified compared with XSA assessment.
- Most figures and tables in the reports were well organized and visualized.


## 9. Technical Comments

- Table 25.2 is not clear.
- It's difficult to read the numbers in Figures 25.11 and 25.12.
- Figure 25.14 is in low resolution and difficult to read.


## 10. Conclusions

- The assessment of Plaice in division 7.e appears well written with no large errors.
- The RG recommends that the assessment is accepted.


## Plaice (Pleuronectes platessa) in divisions 7.f-g (Celtic Sea Plaice) (ple.27.7fg)

1. Assessment Type:

- Benchmark

2. Assessment:

- Accept


## 3. Forecast:

- None.


## 4. Assessment Model:

- WKLIFE: $r f b$ rule is used for advice
- SPiCT was applied for 2022 with high uncertainty and generally rejected for management use.
- SAM (State-Space Assessment) was suggested in the 2022 benchmark assessment. However, it resulted in a strong retrospective pattern with high Mohn's Rho and provided contradictory conclusions. Therefore, the $r f b$ rule was applied.


## 5. Consistency:

- Commercial abundance indices from the different fisheries provided inconsistent trends due to varying discarding practice from 2011 onwards. The situation began to return to normal after 2018 when most fish of commercial size were retained.
- The SPiCT model was used as the basis for advice in 2021. However, the SPiCT model converged with very large uncertainty and with unsatisfactory diagnostics.


## 6. Stock Status:

- The fishing pressure was below $\mathrm{F}_{\text {MSY proxy. }}$. However, the stock size was below $\mathrm{I}_{\text {trigger }}$.
- Total landings from 2021 were equal to 846 t , which was $56.7 \%$ lower than the TAC of 1911 t . The stock has declined to close to the lowest observed from 2016-2021, and the advised TAC was reduced to 1735 t .
- Plaice is primarily a bycatch species in the targeted sole fishery and nephrops fishery. Discards are a large proportion of total catch ( $44.6 \%$ in 2021), and highly variable among years). 2019-2021 was the first time that landings exceeded discards


## 7. Management Plan:

- There is a TAC for Celtic Plaice, and a mean length reference point for a comparison with $\mathrm{MSY}_{\text {proxy }}$ to guide management.
- TAC is divided into national quotas.
- Beyond TAC there is no further management plan.


## 8. General Comments:

- The justification for implementation of the previous modeling methodologies is well presented, giving the rationale for the current modeling methodology and describing the data limitations for this species.
- The abundance decreased significantly since the 2010s, and this stock appeared to make strong year classes occasionally. The RG suggests investigating the relationship between recruitment and environmental conditions as well as examining if the decline in abundance was associated with environmental variability or overfishing.
- The landings weight-at-age have decreased since the 2000s especially for large age groups (age 6+ older). The RG suggests the WG evaluate the impacts of decreased weight-at-age on the productivity of this population and explore the possible reasons for the decline in weight-at-age.
- The plaice stock in 7 fg is considered primarily a bycatch of the targeted sole fishery. Although its ecological importance in the ecosystem, the discards rates have been high due to mismatch of gear selectivities as well as low market value. The fishing mortality on plaice has been highly impacted by changes in fishing effort in the targeted sole fishery. The RG encourages the WG to consider using multispecies assessment methods for this stock.


## 9. Technical Comments:

- Consider putting more contrasting color within Figure 28.1 and zooming into the field space. It is currently difficult to see points with smaller values. Also, how much space does each square on the grid represent?
- Figure 28.1 is missing a general area description or latitude/longitude values to give spatial reference.
- Figure 28.11 missing a legend for associated fishing mortality at age
- Country order of TACs and Quotas table should be consistent
- Figure 28.4 and 28.13, axis labels are small and difficult to read.
- Figure 28.8 is cut off by page format.
- Table 28.5: The unit should be "Numbers $10^{3 "}$ ", spelling errors in figure description.
- Table 28.6: Please specify the term "SOPCOFAC" in the table.
- Table 28.10: Please add column names and row names.


## 10. Conclusions

- The WG report is well-written and organized. The concerns and limitations of available data were documented and discussed in the report. The RG appreciates the WG's efforts to address the issues in the data and historical model configurations.
- The RG recommends the assessment to be accepted.


## Plaice (Pleuronectes platessa) in divisions 7h-k (Celtic Sea South, southwest of Ireland) (ple.27.7hk)

1. Assessment Type:

- Update
- Last Benchmark: 2021 (not found in report, but found in separate document)

2. Assessment:

- Accept with Caveats

3. Forecast:

- None.

4. Assessment Model:

- WKLIFE X: rfb rule.

5. Consistency:

- No stock annex was provided for plaice in divisions $7 \mathrm{~h}-\mathrm{k}$.
- This is the second year that the current methodology is being used.
- Stock identification of plaice is not based on tagging or genetic studies.

6. Stock Status:

- Relative fishing mortality is below the $\mathrm{F}_{\text {MSY }}$ proxy in 2020.
- Relative biomass is above the reference $\mathrm{B}_{\mathrm{MSY}} * 0.5$ proxy in 2020.
- Spawning biomass has remained high and stable since 2004.
- Fishing mortality has been below $\mathrm{F}_{\text {MSY }}$ since 2004, and is now at the lowest point in the time-series.
- Official landings in 2021 (46 tonnes) were $31 \%$ lower than the 2021 TAC (67 tonnes).


## 7. Management Plan:

- 2022 TAC = 114 tonnes
- TAC for 2021 was 67 tonnes
- The management advice for each of the years 2023 and 2024 when the MSY approach is applied:
- Catches should not exceed 132 tonnes.
- Reference points:
- MSY $_{\text {trigger proxy }}=150 \mathrm{mt}$
- $\mathrm{F}_{\text {MSY proxy }}=1$


## 8. General Comments:

- The RG commends the WG on a well written assessment.
- The WG does a very good job explaining the effects that COVID-19 had on sampling and data collection.
- The RG suggests inclusion of more justification on why VAST was applied to combine the indices, rather than using individual indices. Also including when and why the indices started to be combined using VAST for the assessment would be beneficial.
- The WG should include justification on why no temporal autocorrelation was included in VAST. While the residuals do not show concerning patterns, the WG should consider trying multiple models with different formulations, and then comparing their performance to pick the best one. As part of these explorations, inclusion of environmental covariates in the model, especially considering the influence of climate change in the North Atlantic would be beneficial.
- For the calculation of the $r f b$ index, $\mathrm{L}_{\mathrm{C}}$ and $\mathrm{L}_{\text {mean }}$ are calculated as a mean over the time series. Are these estimated from all data over the time series 2004-2021 or as a mean estimated from each year's data? This is an important distinction and justification of which one was used is needed.
- Additionally, given the values of $\mathrm{L}_{\mathrm{C}}$ and $\mathrm{L}_{\text {mean }}$ in the Figures (29.16 and 29.17), it appears that these values have a temporal trend with early time series values being lower and later time series values being higher (in general). Why then is the entire time series 2004-2021 used for these calculations? Temporal shift of these yearly estimations may mean that current values of $\mathrm{L}_{\mathrm{C}}$ and $\mathrm{L}_{\text {mean }}$ are underestimations. More justification is needed.
- The WG should consider including a section detailing the nature of the biomass indices. Consider comparing the individual indices to the combined index, and quantifying uncertainties in the indices.
- The RG recommends including a table in section 29.3.4 that includes all of the calculated values and their summary statistics.
- The WG should include more discussion on why this stock should be considered for SPiCT in section 29.1. Additionally, the WG has no mention of previous exploratory use of SPiCT as is outlined by the WKLIFE X flowchart. If these exploratory assessments were completed, information and any results (even if not used) should be included.
- After communications with the WG, the RG understands the reason for lack of an annex for the current assessment. The RG will not fault the current assessment for this, but requests an annex be made before the next update and requests further
elaboration on previous management and assessment methods to be included in the current report. This will greatly increase understanding and readability in the absence of an annex.


## 9. Technical Comments

- The years of the last update and benchmark assessments should be included in the report.
- Discussion of the landings obligation (section 29.2.2) is not clear and the term "survivability exemption" is not defined.
- There is no discussion of how Ay is calculated. If it is a TAC from a previous year, it does not match the TAC of 2022 (114) or 2021 (67).
- Section 29.4.2 needs rewording for clarification.
- The WG should check the numbering of tables/figures throughout the assessment. For instance, in section 291.3.4, Figure 29.18 is referenced for the mean length in observed catch, but that figure is the length-age relationship.
- The tables in section 29.4.1 and 29.4.3 need table captions.
- The TAC tables in section 29.2.3 need captions, and the units need to be included.
- Generally, all figures have small font sizes and are difficult to read.
- Figure 19.1 is blurry.
- Figures 29.2 and 29.3 color variations between countries lead to poor readability, especially colors chosen for the Netherlands, Spain, and UK (England).
- Figure 29.8 color variation between the surveys leads to poor readability.
- Figures 29.11a and b are blurry and colors are indistinguishable from each other. Additionally, the range of the scales for each plot should be reevaluated.
- Figure 29.12 is missing an X axis label.
- In section 29.3.4, paragraph 2, sentence 7, "was found to be" is repeated twice.
- In section 29.3.4, paragraphs 3 and 4, "von Bertalanffy" is spelled incorrectly.
- Figures should have consistency in colors for labels for different counties.
- Figures 29.16 and 19.17 could use greater description of figures in captions.
- In general, all Figure captions should be expanded for further clarification.


## 10. Conclusions

- The assessment of plaice in divisions $7 \mathrm{~h}-\mathrm{k}$ is well done and well written.
- The RG agrees with the WG and also recommends further research on tagging and genetics in future benchmarks so as to further justify stock boundaries.
- After communications with the WG, the RG understands the reason for lack of an annex for the current assessment. The RG will not fault the current assessment for this, but requests an annex be made before the next update and requests further
elaboration on previous management and assessment methods to be included in the current report.
- The RG recommends the assessment of plaice in divisions h-k to be accepted under the following caveats:
- Discussion of any exploratory SPiCT use is discussed or the lack of use is further justified.
- Because of the lack of an annex, discussion of previous management and assessment practices is necessary for the report.
- The additional general and technical comments are addressed.


## References

Carruthers, T.R., and Hordyk, A. 2018. The Data-Limited Methods Toolkit (DLMtool): An R package for informing management of data-limited populations. Methods Ecol Evol. 2018;1-8.
Carruthers, T.R., Kell, L.T., Butterworth, D.D.S., Maunder, M.N., Geromont, H.F., Walters, C., McAllister, M.K., Hillary, R., Levontin, P., Kitakado, T., and Davies, C.R. 2016. Performance review of simple management procedures. ICES Journal of Marine Science, 73 (2): 464-482.
Carruthers, T.R., Punt, A., Walters, C., MacCall, A., McAllister, M., Dick, E.J., Cope, J., 2014. Evaluating methods for setting catch limits in data-limited fisheries. Fisheries Research, 153: 48-68.
Newman, D., Carruthers, T.R., MacCall, A., Porch, C., Suatoni, L. 2014. Improving the Science and Management of Data-Limited Fisheries: An Evaluation of Current Methods and Recommended Approaches, Natural Resources Defense Council.


[^0]:    ICES INTERNATIONAL COUNCIL FOR THE EXPLORATION OF THE SEA CIEM CONSEIL INTERNATIONAL POUR L'EXPLORATION DE LA MER

[^1]:    * The figures in the table are rounded. Calculations were made with unrounded inputs, and computed values may not match exactly when calculated using the rounded figures in the table.
    ${ }^{* *} A y+1=A y \times r \times f \times b \times m$.
    *** [Advised catch for 2023] $\times$ [1 - discard rate].
    ^ Advice value for 2023 relative to the advice value for 2022 ( $\mathbf{1 4} 116$ tonnes).

[^2]:    Assess- Previous cat1 assessment used FLXSA.Development of a SAM assess-Assessment ment Poor catch sampling leads to substantialment. input data. method uncertainty in the catch at age data used in the assessment - a more appropriate

[^3]:    * Data calculated using estimates from discard observer trips.

[^4]:    *Average weights 2017-2019.

[^5]:    * Discards estimates available since 2007, prior to 2007 discards estimates are based on limited sampling.
    **Geometric mean recruitment 1993-2019.

[^6]:    * Number of trips expressed as number of hauls for discards.

[^7]:    * Based on 2017-2019 discard rates

[^8]:    * Values for 2010, 2011 and 2012 are not reliable due to poor sampling.
    ** Values for 2016 revised at WGCSE 2018 due to inclusion of Northern Irish sampling in 2016 and 2017.
    *** No sampling in 2020; values for 2020 are based on 2017-2019 averages
    **** Only Northern Irish samples; values for 2021 are based on 2019 and 2021 averages

[^9]:    * provisional.

[^10]:    ${ }^{1}$ There is a large price differential between the large and small grades. So less volume of the larger grade generates an economically viable return for fishing.

[^11]:    * reduced isometric grid.
    ** mean density of the observations.

[^12]:    * 2022 abundance estimate used in the assessment was based on 2021 UWTV survey abundance estimate.

[^13]:    * the 2013 survey achieved partial coverage $\sim 60 \%$ of the total area. The abundance has been scaled up to the entire area since densities in the unsurveyed part of the ground were not significantly different in 2014.
    $\mathrm{nr}=$ no reliable abundance estimate could be calculated because survey coverage was partial.
    ** mean density adjusted of the observations.

[^14]:    * Preliminary.
    ${ }^{\text {c }}$ Incomplete/missing due to part of the data being unavailable under national GDPR clauses.

[^15]:    ${ }^{1}$ https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX\%3A32020R0123
    ${ }^{2 h}$ https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX\%3A02021R0092-20211101
    ${ }^{3}$ https://eur-lex.europa.eu/legal-content/FR/ALL/?uri=CELEX:52022PC0054

[^16]:    *Not used in assessment (lack of information. High probability of underestimation considering management measures).

[^17]:    * Preliminary.

[^18]:    ${ }^{1}$ Value corrected. The assessment and advice were revised due to the discovery of an error in the setting of the catch numbers for age 1 and 2 for 1971-2003 in the SAM model. Values and figures have been updated in this sheet.

[^19]:    ${ }^{2}$ TMB offers a modelling framework for fast estimation of hierarchical models written in C code through the Laplace approximation. In addition, increased performance of nonlinear optimization procedures is achieved through the use of AUTODIFF (automatic differentiation), and performant C libraries for linear algebra (Eigen and CholMod).

[^20]:    * EU multiannual plan (MAP) for the Western Waters (EU, 2019).

[^21]:    * The number of sampled trips that took place.

[^22]:    *this includes the ICES estimate of $102 \mathbf{t}$ for IR-OTB discards.

[^23]:    ${ }^{1}$ https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX:32019R2239

[^24]:    ${ }^{2}$ https://community.ices.dk/ExpertGroups/benchmarks/2019/wkceltic/2014\%20Meet-
    ing\%20docs/02.\%20Background\%20documents/WHG/aggregate_IC data_whg.27.7b-ce-
    k Oct 2020.html

[^25]:    3
    https://community.ices.dk/ExpertGroups/benchmarks/2019/wkceltic/2014\%20Meeting\%20docs/04.\%20Working\%20documents/WD 03 WKCELTIC\%20-\%20French\%20commercial\%20tuning\%20fleets Final 2020.pdf
    ${ }^{4} h \mathrm{https}: / /$ community.ices.dk/ExpertGroups/wgcse/2022\%20Meeting\%20Documents/06.\%20Data/whg-7b-ce-k/aggregate_IC data_whg. $27.7 \mathrm{~b}-\mathrm{ce}-\mathrm{k}$ Sept 2022.html

