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## 5 Cod in 7.a (Irish Sea)

Situated between Ireland and Great Britain the Irish Sea (7.a) is connected by to the Celtic Sea (7.g) at its southern extreme by the St George's Channel and in north is linked to sea region West of Scotland (6.a) by the Northern Channel. The average depth is 50 m but the area is contrasted between a deeper channel, in the west, and shallower bays in the east. The channel has a maximum depth exceeding 275 m whilst the eastern bays have depths less than 50 m . Distinct habitat patches result from a combination of bathymetry, topographical features and hydrography. The sea bed of the eastern Irish Sea is dominated by fine sediment plains with some small areas of areas of mud habitat, the fine sediments graduate to more coarse material in central areas. A large well-defined deep-water mud basin is located in the northwestern region in close to the Northern Irish and Irish coast.

Irish Sea fisheries are predominantly demersal trawling and seining with demersal trawling for Nephrops dominating effort with vessels using mesh in the range $70-99 \mathrm{~mm}$. Effort using fishing gear with $\geq 100 \mathrm{~mm}$ mesh sizes is currently at a low level compared to historic activity, a considerable decline in effort was observed between 2003 and 2007 and has continued since. The species composition of catches by vessels in using $\geq 100 \mathrm{~mm}$ mesh consists primarily of haddock, with lower quantities of hake. At present there is no commercial towed gear fishery for cod permitted. Beam trawls are operating within the Irish Sea with mesh sizes in the range $80-119 \mathrm{~mm}$, targeting sole, plaice, and rays. A seasonal pelagic and gillnet herring fishery operates in late summerearly autumn in the pre and post-spawning period. Dredge fisheries target king and queen scallops, with king scallops in coastal areas with the queen scallop fishery operating in the central area south of the Isle of Man, to a lesser extent queen scallops are also targeted using trawl nets, during the late summer when swimming activity is most pronounced.
There is a recreational fishery which catches cod and with declining commercial rates has become a more important aspect of the total catch. At the last benchmark in February 2022 (ICES, 2022a) the recreational fishery was included in the assessment for the first time.

## Type of assessment

The stock was benchmarked in February 2022 (ICES, 2022a) and a Stock Synthesis (SS3) fully analytical model is now being used in the cod assessment.

## ICES advice applicable to 2021 and 2022

ICES advised on the basis of precautionary approaches that there should be no directed fisheries, and bycatch and discards should be minimized in 2021 and 2022. Advice since 2020 was applied based on the on the 2 over 3 rule for category 3 .

## ICES advice applicable to 2023

ICES advised on the basis of the MSY and precautionary approach that there should be zero catches in 2023 as SSB will be below Blim in 2023 and 2024.

### 5.1 General

## Stock description and management units

The stock and the management unit are both ICES Division 7.a (Irish Sea).

## Management applicable to 2021

TACs and quotas set for 2021

|  | TAC | Landed |
| :--- | :---: | :---: |
| Belgium | 3 | 2.8 |
| France | 7 | 0 |
| Ireland | 104 | 41.8 |
| The Netherlands | 1 | 0 |
| United Kingdom | 91 | 88.7 |
| EU (total) | 115 | 44.6 |
| Total | 206 | 133.3 |

Management of this cod fishery is by TAC, days-at-sea limits and technical measures. Technical regulations in force in the Irish Sea, including those associated with the cod recovery plan since 2000, are described in Section 5.2.

Quota uptake in 2021 was considerably below the officially set TAC of 206 tonnes.
Table 1. Fishing opportunities (TAC) for 2022 for cod in 27.7.a.

| Species: | Cod Gadus morhua |  | Zone: | 7a (COD/07A.) |
| :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 | (1) | Precautionary TAC <br> Article 7(1) of this Regulation applies |  |
| France | 7 | (1) |  |  |
| Ireland | 104 | (1) |  |  |
| Netherlands | 1 | (1) |  |  |
| Union | 115 | (1) |  |  |
| United Kingdom | 91 | (1) |  |  |
| TAC | 206 | (1) |  |  |

(1) Exclusively for bycatches. No directed fisheries are permitted under this quota.
https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX\%3A32021R1239.

## Fishery in 2021

Landings in accordance with TAC were below the TAC, however, the TAC in 2021 was considerably higher than the ICES advised value of 74 tonnes.

Since 2009, Irish landings of cod reported from ICES rectangles immediately north of the Irish Sea/Celtic Sea boundary (ICES rectangles 33E2 and 33E3) have been reallocated into the Celtic Sea as they represent a combination of inaccurate area reporting and catches of cod considered by ICES to be part of the Celtic Sea stock (ICES, 2009). The amount of Irish landings transferred
from 7 a to $7 \mathrm{e}-\mathrm{k}$ by year is shown below. Total official landings for this stock in 2021 were 133 tonnes after this re-allocation and total catches in the area were 184 t .

Table 2. Transfers from ICES rectangles 33E2 and 33E3.

| Year | Tonnes |  |
| :---: | :---: | :---: |
|  | 2004 | 108 |
|  | 2005 | 54 |
|  | 2006 | 103 |
|  | 2007 | 527 |
|  | 2008 | 558 |
|  | 2009 | 193 |
|  | 2010 | 143 |
|  | 2011 | 147 |
|  | 2012 | 130 |
|  | 2013 | 75 |
|  | 2014 | 24 |
|  | 2015 | 39 |
|  | 2016 | 40 |
|  | 2017 | 19 |
|  | 2018 | 20 |
|  | 2019 | 37 |
|  | 2020 | 71 |
|  | 2021 | 52 |

The majority of landings in 2021 was taken by the Nephrops fleet, followed by the midwater demersal fleet. Landings and discards by métier and country can be seen in Table 8. Total uptake of cod TAC was $65 \%$.

A Fishery-Science Partnership Survey (FSP) was repeated in the western Irish Sea in spring 2021 in the western Irish Sea using semi-pelagic gear on commercial vessels. This survey attempts to address the lack of sampling opportunities created by the diminishing TAC for cod in the Irish Sea and the resulting significant reduction of a directed whitefish fleet targeting cod.

## InterCatch procedure

Since 2013 international landings and discards-at-age are uploaded into InterCatch. Discards are raised for unreported strata and métiers to estimate total discards-at-age.

## Landings

The input data on fishery landings and age compositions are split into four periods:

1. 1968-1990. Landings in this period, provided to ICES by stock coordinators from all countries, are assumed to be un-biased and are used directly as the input data to stock assessments.
2. 1991-1999. TAC reductions in this period caused substantial misreporting of cod landings into several major ports in one country, mainly species misreporting. Landings into these ports were estimated based on observations of cod landings by different fleet sectors during regular port visits. For other national landings, the WG figures provided to ICES stock coordinators were used.
3. 2000-2005. Cod recovery measures were considered to have caused significant problems with estimation of landings. The ICES WG landings data provided by stock coordinators for all countries are considered uncertain and estimated within an assessment model. Observations of misreported landings were available for 2000, 2001, 2002 and 2005. However, they have generally not been used to correct the reported landings but have been used to evaluate model estimates in those years.
4. Since 2006. The introduction of the UK buyers and sellers legislation is considered to have reduced the bias in the landings data but the level to which this has occurred is unknown. Consequently comparisons were made between the fit of the model to recorded landings under an assumption of bias and unbiased information.
5. 2020. The Covid-19 pandemic made the collection of observer data aboard vessels impossible for Q2-Q4, making the estimation of discard data and the establishment of age structure in catches impossible for most of the year. Age structure of the stock is available from Q1 observer data and the 3 surveys, FSP, and Q1 and Q4 groundfish surveys.
1. 2021. The continued COVID-19 situation resulted in reduced sampling; for the quarter 1 2021 the full final tow of the TR1 fleet was landed and sampled by observers ashore. There was very low sampling of cod in the Nephrops directed fleet, particularly in quarter 1 due to no observed trips. A raising procedure similar to the previous year was applied, in which the cod sampled in the Northern Ireland fishers self-sampling scheme were applied and raised to the full Nephrops catches. However, no cod were found in the provided self samples.

The annual numbers-at-age caught and the mean weights-at-age in landings (applied to the total catch) by age are given in Table 9 and Table 10; numbers of catch-at-age for 2020 are excluded due to limited discard and port sampling during the COVID-19 pandemic.

## Discards data

The WKIrish3 (ICES, $2017 \mathrm{a}, \mathrm{b}$ ) benchmark report gives details on historic raising to total national and international discards.

## Biological data

## Natural mortality

Natural mortality has been revised in WKNSCS (ICES, 2022a). M-at-age was calculated from tagging data following calculated following (Pollock, Hoenig et al., 1989, Hoenig, Barrowman et al., 1998). Natural mortality is kept constant throughout years.

## Maturity

Maturity ogive has been revised in WKIrish2 (ICES, 2016). Each year the smoother is applied to the full time-series of raw data and values are accordingly updated. Updated values after application of the smoother are in Table 12. Please refer to the stock annex for further information.

## Survey data used for advice

Please refer to the stock annex for a description of the surveys and survey data.

| Survey | Ages | Years |
| :--- | :---: | :---: |
| FSP SURVEY (B7897) | $2-6$ | $2004-2021$ (EXCLUDING 2014) |
| NIGFS-WIBTS-Q4 (G7655) | 0 | $1995-2021$ |
| NIGFS-WIBTS-Q1 (G7144) | $1-4$ | $1995-2021$ |

### 5.2 Historical stock development

The advice is based on the newly benchmarked assessment (WKNSCS, ICES 2022a).

## Deviations from Stock Annex

During the benchmark process kept and assumed dead (i.e. $35 \%$ of released fish) removals from the recreational fishery were included in the assessment. There was considerable discussion regarding the introduction of the recreational fishery, which was in the range of 30 tonnes for each of the available data year 2017-2020. The benchmark agreed and reviewed a) a model excluding the recreational data due to issues with uncertainty of data and insufficient information of a selectivity pattern and b) combine the recreational removals with the total commercial catches, using the same selectivity pattern. With a view on a possible increase of the recreational component over the next few years with decreasing commercial fishing pressure, the benchmark decided that it would be beneficial to go for approach $b$ to enable the incorporation of future recreational removals. The differences in SSB, Fbar and general perception of the stock were negligible.

In preparation for the working-group it became apparent that the values estimated for the recreational fleet ahead of the benchmark contained a data error and are now indeed estimated at approximately 120-150 tonnes for each of the year 2017-2021, similar levels to the commercial catches.

There are high uncertainties around the recreational removals and the values largely diverge from the benchmarked values.

The benchmark had agreed on a model excluding the recreational catches and passed the model as fit to benchmark and considered the inclusion of the recreational data only as a way to include them in times of increase in relation to commercial catches. In light of the data error the WGCSE working group therefore decided in this instance to exclude all recreational removals from the model.

## Final assessment

The final assessment has been run in stock Synthesis (SS3). Available data and catch-at-age, dis-cards-at-age and numbers-at-age in surveys can be seen in Figure 1-3, Figure 6 and Table 9-13, while summary of assessment results can be seen in Table 14.

The fit of the model catch-at-age data and to the indices is good, showing "all green" runs tests (Figure 11 and Figure 12) as well as the individual residuals. Further details on the use of the Runs tests and RMSE can be found in Carvalho et al., 2021. The retrospectives provide a good fit with Mohns Rho for SSB and $\mathrm{F}_{\mathrm{bar}}$ at 0.09 and -0.14 respectively (Figure 13).

The final results of the assessment can be seen in Figure 4-6.

## Final assessment: long-term trends

### 5.3 Short-term predictions

Short-term forecast was carried out in using the FLR forecast environment. Assumptions for the intermediate year can be seen in Table 3. Geometric mean for recruitment is from 2002-2019 (Final year-2), which encompasses the block where recruitment is supposedly reduced.

Table 3. Short-term forecast assumptions.

| Variable | Value | Notes |
| :--- | :---: | :--- |
| $F_{\text {ages 2-4 (2022) }}$ | 0.038 | $F_{\text {sq }}=F_{\text {average (2018-2021) }} *$ |
| SSB (2023) | 4842 | Short-term forecast fishing at $f_{\text {sq; in tonnes. }}$ |
| $R_{\text {age } 0}(2022$ and 2023) | 17989 | Geometric Mean (2002-2019); in thousands |
| Total catch (2022) | 165 | Fishing at $F_{\text {sq; in tonnes }}$ |
| Projected landings (2022) <br> ((2022)((20(2022(2020) | 159 | Assuming average landing patterns (2019-2021); in <br> tonnes |
| Projected discards (2022) | 6 | Assuming average discard patterns (2019-2021); in <br> tonnes |

* $F$ in 2020 was assumed to be unrepresentative due to the COVID-19 disruption and hence $F_{s q}$ was calculated as $F_{\text {average }}$ (2018-2021) excluding 2020.

Table 4 shows the catch scenarios, in particular the zero catch advice and the scaled MSY advices due to SSB being below MSY Btrigger and unable to reach Blim even under a no-catch scenario. The newly introduced Feco (ICES, 2022a) is also included in a scaled version. With the Sea Surface temperature Index being being high for the recent years, FECO is currently set at 0.19.

Table 4. Catch scenarios for 2023; all weights are in tonnes.

| Basis | Total catch (2023) | Projected landings (2023) | Pro- <br> jected <br> dis- <br> cards <br> (2023) | $\begin{aligned} & F_{\text {total }} \\ & (2023) \end{aligned}$ | $F_{\text {projected }}$ <br> landings <br> (2023) | $F_{\text {projected dis- }}$ cards (2023) | $\begin{aligned} & \text { SSB } \\ & \text { (2024) } \end{aligned}$ | \% SSB change | \% TAC change $\wedge$ | \% Advice change ${ }^{\wedge \wedge}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ICES advice basis |  |  |  |  |  |  |  |  |  |  |
| MSY approach: $F=0$ | 0 | 0 | 0 | 0 | 0 | 0 | 5410 | 11.7 | -100 | -100 |
| Other scenarios |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \mathrm{F}_{\text {MSY }} \times \text { SSB } \\ & (2023) / \text { MSY } \text { Brrig. }^{\text {ger }} \\ & \text { g. } \end{aligned}$ | 403 | 382 | 21 | 0.093 | 0.090 | 0.0036 | 4988 | 3.0 | 96 | 440 |
| $\begin{aligned} & \text { FMSY lower } \times \text { SSB } \\ & (2023) / \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |
| $F=\mathrm{F}_{\text {MSY }}$ lower | 704 | 667 | 37 | 0.168 | 0.162 | 0.0065 | 4677 | -3.4 | 240 | 850 |
| $F=F_{\text {MSY }}$ | 908 | 861 | 48 | 0.22 | 0.21 | 0.0086 | 4466 | -7.8 | 340 | 1130 |
| $F=F_{p a}$ | 1011 | 958 | 53 | 0.25 | 0.24 | 0.0097 | 4362 | -9.9 | 390 | 1270 |
| $\mathrm{F}=\mathrm{F}_{\text {MSY }}$ upper | 1093 | 1035 | 58 | 0.27 | 0.26 | 0.0106 | 4278 | -11.7 | 430 | 1380 |
| $F=F_{2022}$ | 170 | 161 | 9 | 0.038 | 0.037 | 0.00148 | 5232 | 8.0 | -17.5 | 130 |
| $\mathrm{F}=\mathrm{F}_{\text {lim }}$ | 1612 | 1526 | 86 | 0.43 | 0.41 | 0.0166 | 3754 | -22 | 680 | 2100 |
| $\mathrm{F}=\mathrm{F}_{\text {Eco }}$ | 788 | 747 | 41 | 0.19 | 0.183 | 0.0074 | 4589 | -5.2 | 280 | 960 |
| $\begin{aligned} & \mathrm{F}_{\mathrm{ECO}} \times \operatorname{SSB} \text { (2023) } \\ & / \mathrm{MSY} \mathrm{~B}_{\text {trigger }} \end{aligned}$ | 347 | 329 | 18 | 0.080 | 0.077 | 0.0031 | 5047 | 4.2 | 68 | 370 |
| $\begin{aligned} & \text { SSB (2024) = } \\ & \text { SSB (2023) } \end{aligned}$ | 544 | 516 | 28 | 0.128 | 0.123 | 0.0049 | 4842 | 0 | 160 | 640 |
| $\underset{* *}{\operatorname{SSB}(2024)}=\text { Blim }^{2}$ |  |  |  |  |  |  |  |  |  |  |

* SSB 2024 relative to SSB 2023.
** The Blim option was left blank because Blim cannot be achieved in 2024, even with zero catches.
${ }^{\wedge}$ Total TAC in 2023 relative to the TAC in 2022 ( 206 tonnes).
$\wedge \wedge$ Total Advice in 2023 relative to advice in 2022 ( 74 tonnes).


### 5.4 Biological reference points

New reference points were defined at WKNSCS (ICES, 2022a). The newly introduced Feco (ICES, 2022a) has been agreed and reviewed at the benchmark for a stock for the first time. FECO is an opportunity to use environmental data in forecast scenarios (ICES, 2022a). In case of cod in 7 .a a sea surface temperature (SST) was found to be a reasonable indicator for productivity. The Feco reference point uses the inverted SST (with a 3-year lag to account for the time from larvae stage to contribution to SSB) rescaled between zero and one which informs the status of the indicator
(Is) in the advice year compared with previous years. The status of the indicator determines the placement of the FECO reference point within FMSY ranges (ICES, 2019; 2020); for 2023 FECO is at 0.19 , estimated as FMSY lower + ((FMSY upper-FMSY lower)*Is).

Table 5. Biological reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $B_{\text {triger }}$ | 11538 | $\mathrm{B}_{\mathrm{pa}}$ | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.222 | Median point estimates of ( $\mathrm{F}_{\text {MSY }}$ ) EqSim with combined SR | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\text {MSY lower }}$ | 0.168 | Median lower estimates of ( $\mathrm{F}_{\text {MSY }}$ ) EqSim with combined SR | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\text {MSY upper }}$ | 0.273 | Median upper point estimates of ( $\mathrm{F}_{\text {MSY }}$ ) EqSim with combined SR | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\text {ECO }}$ | 0.19 | $\begin{aligned} & \text { Ecosystem Indicator }\left(I_{s}\right) ; \mathrm{F}_{\mathrm{ECO}}=\mathrm{F}_{\mathrm{MSY} \text { lower }}+\left(\left(\mathrm{F}_{\mathrm{MSY} \text { upper }}-\right.\right. \\ & \left.\left.\mathrm{F}_{\text {MSY lower }}\right)^{*} I_{s}\right) \end{aligned}$ | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
| Precautionary approach | $\mathrm{Bl}_{\text {lim }}$ | 8303 | Lowest SSB with above-average recruitment | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 11538 | Blim combined with the assessment error | ICES, 2022a |
|  | $\mathrm{F}_{\text {lim }}$ | 0.43 | F with 50\% probability of SSB less than $\mathrm{Bl}_{\text {lim }}$ | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.25 | $\mathrm{F}_{\text {P05 }}$; the F that leads to $\mathrm{SSB}>\mathrm{B}_{\text {lim }}$ with $95 \%$ probability | $\begin{aligned} & \text { ICES, } \\ & \text { 2022a } \end{aligned}$ |

### 5.5 Management plans

The Irish Sea cod management plan, as described in Council Regulation (EC) 1342/2008 was evaluated independently by ICES in 2009 using the approach adopted in AGCREMP 2008 and found to be not consistent with the ICES Precautionary Approach (WGCSE 2009).

### 5.6 Uncertainties and bias in assessment

## Surveys

The Irish Sea has relatively good survey coverage. The quarter 1 groundfish survey and the FSP survey have got good consistent cover of the age contributions. The Q 4 groundfish survey only attributes to the recruitment at age 0 .

## Stock structure and migrations

Stock structure and migrations have been in full discussed in the WKIrish2 report (ICES, 2016), however, there are still uncertainties and discussions.

A tagging study of Irish Sea cod and Celtic Sea cod was conducted from 2016-2019 in part to address these issues. Up to January 2019, 4238 cod were caught and tagged aboard chartered commercial fishing vessel using semi-pelagic fishing gear, FSP survey, shore angling competitions and others. Up to January 2019, 138 tagged cod were returned. The project relies on collaboration with the fishing industry to provide the data to develop a better understanding of the current behaviour, biology and stock status of Irish Sea cod. Most recent results suggest a stronger migratory behaviour of Irish Sea cod into the Celtic Sea, indicating that up to $18 \%$ of mature fish might leave the Irish Sea (ICES, 2021). This will have considerable impacts on the future management and assessment of the stock, but additional research is necessary. Currently a further project using data storage tags and trace element analysis is being conducted to understand stock structure and migratory behaviour as well as mixing.

### 5.7 Management considerations

A number of emergency and cod recovery plan measures have been introduced since 2000 to conserve Irish Sea cod. These include a spawning closure since 2000 and effort control since 2003. There have also been several vessel decommissioning schemes. As it has not been possible to provide analytical catch forecasts in recent years, the TAC has been reduced by 15-20\% annually since 2006 and by $25 \%$ since 2009. An MSY approach was used to set TAC in 2018 and 2019, which was followed by a precautionary advice since 2020. Since 2022 the stock is being assessed using an MSY approach; however, low SSB and the incapability of reaching Blim by 2024 even under with zero catches lead to a zero catch advice for 2023.

### 5.8 Future Issues and considerations

Cod in the Irish Sea and the Celtic Sea are in a highly exploited state and show historically a very steep age-profile. Recruitment since 2002 has been impeded.

It is essential to further the understanding of the stock structure to improve future management, which includes the further investigation of migration and natural mortality in the Irish Sea. It might be necessary for a combined approach to manage the stocks in 7.a and 7.e-g.

Under the current highly exploited status it seems that recruitment rather than fishing pressure is driving stock trends. It is also questionable in how far an MSY approach with reference points as applied in the traditional ICES format is a valid approach for this stock which is recruitment rather than fishery controlled. The working group is awaiting the outcomes of WKREF to further investigate the most appropriate way to manage the stock in the future. This might mean a shift to an MSE approach for management.

### 5.9 References

Armstrong, M. J., Gerritsen, H. D., Allen, M., McCurdy, W. J. and Peel, J. A. D. 2004. Variability in maturity and growth in a heavily exploited stock: cod (Gadus morhua L.) in the Irish Sea. ICES J. Mar. Sci., 61, 98-112.
Bendall, V. O., Ó' Cuaig, M., Schon, P-J., Hetherington, S., Armstrong, M., Graham, N., Righton, D. 2009. Spatio-temporal dynamics of Atlantic cod (Gadus morhua) in the Irish and Celtic Sea: results from a collaborative tagging programme. ICES Document CM 2009/J: 06.35 pp .

Carvalho, F.,Winker, H. Courtney, D., Kapur, M., Kell, L., Cardinale, M., Schirripa, M., Kitakado, T., Yemane, D., Piner, K.R., Maunder, M.N., Taylor, I., Wetzel, C.R., Doering, K., Johnson, K.F.,Methot, R.D. 2021. A cookbook for using model diagnostics in integrated stock assessments. Fisheries Research 240.

EU. 2019. Regulation (EU) 2019/472 of the European Parliament and of the Council of 19 March 2019 establishing a multiannual plan for stocks fished in the Western Waters and adjacent waters, and for fisheries exploiting those stocks, amending Regulations (EU) 2016/1139 and (EU) 2018/973, and repealing Council Regulations (EC) No 811/2004, (EC) No 2166/2005, (EC) No 388/2006, (EC) No 509/2007 and (EC) No 1300/2008. Official Journal of the European Union, L 83: 1-17. http://data.europa.eu/eli/reg/2019/472/oj

Hoenig, J. M., et al. 1998. "Multiyear tagging studies incorporating fishing effort data." Canadian Journal of Fisheries and Aquatic Sciences 55(6): 1466-1476.

ICES. 2017. Report of the Benchmark Workshop on the Irish Sea Ecosystem (WKIrish3), 30 January-3 February 2017, Galway, Ireland. ICES CM 2017/BSG:01. 165 pp.
ICES. 2019. Working Group on Multispecies Assessment Methods (WGSAM). ICES Scientific Reports. 1:91. 320 pp. http://doi.org/10.17895/ices.pub.5758.

ICES. 2020. Workshop on an Ecosystem Based Approach to Fishery Management for the Irish Sea (WKIrish6; outputs from 2019 meeting). ICES Scientific Reports. 2:4. 32 pp . http://doi.org/10.17895/ices.pub.5551.
ICES. 2021a. Advice on fishing opportunities. In Report of the ICES Advisory Committee, 2021. ICES Advice 2021, section 1.1.1. https://doi.org/10.17895/ices.advice. 7720 .

ICES. 2022a. Benchmark Report WKNSCS (WKNSCS). Draft report. ICES Scientific Reports.
Pollock, K. H., et al. 1989. Estimation of fishing and natural mortality when a tagging study is combined with a creel survey or port sampling, North Carolina State University. Dept. of Statistics.

Table 6. Official landings ( t ) of COD in Division 7.a as officially reported to ICES and figures used by ICES from 1996. All weights are in tonnes, minor differences in total value are due to rounding. Countries reported landings are official values.

|  |  | 凹 | O ¢ N O |  |  |  |  |  | $\stackrel{\text { ®0 }}{\square}$ |  |  | 出 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 142 | 148 | 2476 | 25 | - | 2359 | 27 | 126 | 5303 |  | 4964** |  |
| 1997 | 183 | 268 | 1492 | 29 | - | 2370 | 19 | 80 | 4441 |  | 5859** |  |
| 1998 | 316 | 269 | 1739 | 20 | - | 2517 | 34 | 67 | 4962 |  | 5318** |  |
| 1999 | 150 | $\mathrm{n} / \mathrm{a}$ | 966 | 5 | - | 1665 | 9 | 80 | 2875 |  | 4784** |  |
| 2000 | 60 | 53 | 455 | 1 | - | 799 | 11 | 38 | 1417 |  | 1274 |  |
| 2001 | 283 | 74 | 751 | - | - | 885 | 1 | 32 | 2026 |  | 2252 |  |
| 2002 | 318 | 116 | 1111 | - | - | 1134 | 7 | 29 | 2715 |  | 2695 |  |
| 2003 | 183 | 151 | 594 | - | 14 | 505 | 7 | 23 | 1477 |  | 1285 |  |
| 2004 | 104 | 29 | 380 | - | - | 646 | 5 | 15 | 1179 | 108 | 1072 |  |
| 2005 | 115 | 35 | 220 | - | - | 594 | n/a | 3 | 967 | 54 | 910 |  |
| 2006 | 60 | 18** | 275 | - | - | 589 | n/a | 6 | 948 | 103 | 840 |  |
| 2007 | 67 | 17** | 608 | - | - | 423 | n/a | 2 | 1117 | 527 | 702 | 148 |
| 2008 | 26 | 3 | 618** | - | - | 543 | 22 | 12 | 1224 | 558 | 661 | 62 |
| 2009 | 19 | 12 | 323** | - | - | 387 | 12 | 12 | 765 | 193 | 468 | 60 |
| 2010 | 21 | 1 | 289 | - | - | 282 | 1 | - | 594 | 143 | 464 | 377 |
| 2011 | 36 | 3 | 275 | - | - | 169 | 1 | - | 485 | 147 | 368 | 43 |


|  | $\stackrel{E}{\square}$ | U | O ¢ ¢ O |  | $\begin{aligned} & \text { 드주 } \\ & \text { in } \end{aligned}$ |  |  |  | $\stackrel{\bar{\square}}{\square}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 23 | 1 | 193 | - | - | 109 | <1 | - | 326 | 85 | 198 | 658 |
| 2013 | 13 | <1 | 160 |  |  | 107 | $<1$ | - | 281 | 76 | 206 | 118 |
| 2014 | 9 | <1 | 148 | - | - | 79 | $<1$ | - | 236 | 24 | 213 | 149 |
| 2015 | 12 | <1 | 137 | - | - | 50 | $<1$ | - | 199 | 39 | 161 | 224 |
| 2016 | 3 | <1 | 84 | - | - | 35 | $<1$ | - | 122 | 40 | 82 | 60 |
| 2017 | 5 | < 1 | 57 | - | - | 41 | <1 | <1 | 103 | 19 | 84 | 59 |
| 2018 | 2 | <1 | 105 | - | - | 128 | $<1$ | <1 | 235 | 20 | 215 | 42 |
| 2019 | 10 | < 1 | - C | - | - | 195 | <1 | <1 | 205 c | 37 | 295 | 7 |
| 2020* | 10 | 0 | 76 | - | - | 95 | <1 | <1 | 252 | 71 | 181 | 25 |
| 2021* | 3 | 0 | 93 | - | - | 89 | $<1$ | <1 | 184 | 52 | 133 | 4 |

* Preliminary official landings.
** Includes sample-based estimates of landings into ports.
*** Landings in the southern part of Division 7.a (rectangles 33 E 2 and 33E3) are not included in the assessment and are considered to be part of the cod stock in divisions 7.e-k.
c Incomplete/missing due to part of the data being unavailable under national GDPR clauses.

Table 7. Working Group figures for annual landings and TAC uptake since 2000. a) total, b) by country.
a)

| Year | Total | TAC | \% uptake |
| :---: | :---: | :---: | :---: |
| 2000 | 1273 | 2100 | 61 |
| 2001 | 2251 | 2100 | 107 |
| 2002 | 2695 | 3200 | 84 |
| 2003 | 1285 | 1950 | 66 |
| 2004 | 1072 | 2150 | 50 |
| 2005 | 910 | 2150 | 42 |
| 2006 | 840 | 1828 | 46 |
| 2007 | 702 | 1462 | 48 |
| 2008 | 662 | 1199 | 55 |
| 2009 | 468 | 899 | 52 |
| 2010 | 465 | 674 | 69 |
| 2011 | 368 | 506 | 73 |
| 2012 | 198 | 380 | 52 |
| 2013 | 206 | 285 | 72 |
| 2014 | 213 | 182 | 117 |
| 2015 | 161 | 146 | 110 |
| 2016 | 82 | 146 | 56 |
| 2017 | 84 | 146 | 57 |
| 2018 | 215 | 695 | 31 |
| 2019 | 298 | 807 | 37 |
| 2020 | 181 | 257 | 70 |
| 2021 | 133 | 206 | 65 |

b)

| 2009 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 391 | 55 | 3 | 19 | 0 | 498 |
| TAC | 259 | 592 | 33 | 12 | 3 | 899 |
| $\%$ uptake | $151 \%$ | $9 \%$ | $9 \%$ | $160 \%$ | $0 \%$ |  |


| 2010 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 292 | 151 | 1 | 21 | 0 | 465 |
| TAC | 194 | 444 | 25 | 9 | 2 | 674 |
| \% uptake | $150 \%$ | $34 \%$ | $4 \%$ | $233 \%$ | $0 \%$ |  |


| 2011 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 170 | 160 | 3 | 36 | 0 | 369 |
| TAC | 146 | 333 | 19 | 7 | 2 | 506 |
| \% uptake | $117 \%$ | $48 \%$ | $16 \%$ | $533 \%$ | $0 \%$ |  |


| 2012 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 112 | 63 | 0 | 23 | 0 | 198 |
| TAC | 109 | 251 | 14 | 5 | 1 | 380 |
| $\%$ uptake | $103 \%$ | $25 \%$ | $0 \%$ | $460 \%$ | $0 \%$ |  |


| 2013 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 107 | 85 | 1 | 13 | 0 | 206 |
| TAC | 82 | 188 | 10 | 4 | 1 | 285 |
| $\%$ uptake | $130 \%$ | $45 \%$ | $10 \%$ | $325 \%$ | $0 \%$ |  |


| 2014 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 79 | 124 | 0 | 9 | 0 | 213 |
| TAC | 52 | 120 | 7 | 2 | 2 | 182 |
| \% uptake | $153 \%$ | $103 \%$ | $0 \%$ | $455 \%$ | $0 \%$ |  |


| 2015 | UK | Ireland | France | Belgium | Netherlands | Total |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 50 | 99 | 0 | 12 | 0 | 161 |
| TAC | 42 | 97 | 5 | 2 | 0 | 146 |
| \% uptake | $119 \%$ | $102 \%$ | $0 \%$ | $600 \%$ | NA |  |


| 2016 | UK | Ireland | France | Belgium | Netherlands | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 35 | 44 | 0.4 | 3 | 0 | 82 |
| TAC | 42 | 97 | 5 | 2 | 0 | 146 |
| \% uptake | 83\% | 45\% | 8\% | 150\% | 0\% |  |
| 2017 | UK | Ireland | France | Belgium | Netherlands | Total |
| Landings | 41 | 38 | 0.2 | 5 | 0 | 84 |
| TAC | 42 | 97 | 5 | 2 | 0 | 146 |
| \% uptake | 98\% | 39\% | 4\% | 250\% | 0\% |  |
| 2018 | UK | Ireland | France | Belgium | Netherlands | Total |
| Landings | 128.5 | 84.6 | 0.05 | 1.9 | 0 | 214.9 |
| TAC | 200 | 459 | 25 | 9 | 2 | 695 |
| \% uptake | 64\% | 18\% | <1\% | <1\% | 0\% | 31\% |


| 2019 | UK | Ireland | France | Belgium | Netherlands | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landings | 193.9 | 90 | 0.2 | 10.2 | 0 | 294.6 |
| TAC | 233 | 530 | 30 | 11 | 3 | 807 |
| \% uptake | 83\% | 17\% | <1\% | 93\% | 0\% | 36.5\% |
| 2020 | UK | Ireland | France | Belgium | Netherlands | Total |
| Landings | 95.6 | 75.9 | 0 | 9.5 | 0 | 181.1 |
| TAC | 74 | 170 | 9 | 3 | 1 | 257 |
| \% uptake | 129\% | 45\% | 0\% | 317\% | 0\% | 70\% |
| 2021 | UK | Ireland | France | Belgium | Netherlands | Total |
| Landings | 88.7 | 41.8 | 0 | 2.8 | 0 | 133.3 |
| TAC | 91 | 104 | 7 | 3 | 1 | 206 |
| \%uptake | 97\% | 40\% | 0\% | 93\% | 0 | 65\% |

Table 8. Landings and discard proportions by métier.

| Catch (2021) | Landings |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 137 tonnes | otter trawls |  | midwater trawl | beam trawls | other gear types |
|  | Nephrops directed | demersal fish directed | 18.7\% | 9.4\% | 1\% |
|  | 33\% | 37\% |  |  |  |
|  | 133 tonnes |  |  |  |  |
|  | Discards |  |  |  |  |
|  | otter trawls |  | midwater trawl | beam trawls | other gear types |
|  | 77\% Nephrops directed | < $1 \%$ demersal fish directed | <1\% | 22\% | 1\% |
|  | 4 tonnes |  |  |  |  |

Table 9. Total catch numbers-at-age (thousands).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 17 | 439 | 1563 | 1003 | 456 | 177 | 30 |
| 1969 | 20 | 969 | 1481 | 1050 | 269 | 186 | 113 |
| 1970 | 22 | 1810 | 1385 | 352 | 204 | 163 | 71 |
| 1971 | 22 | 2835 | 2022 | 904 | 144 | 67 | 51 |
| 1972 | 26 | 900 | 3267 | 824 | 250 | 58 | 59 |
| 1973 | 27 | 2377 | 1091 | 1783 | 430 | 173 | 81 |
| 1974 | 16 | 601 | 3559 | 557 | 494 | 131 | 74 |
| 1975 | 26 | 1810 | 642 | 1407 | 294 | 249 | 117 |
| 1976 | 27 | 1247 | 3007 | 363 | 500 | 61 | 104 |
| 1977 | 31 | 946 | 511 | 1233 | 163 | 218 | 71 |
| 1978 | 40 | 855 | 1092 | 310 | 311 | 39 | 65 |
| 1979 | 44 | 1948 | 1288 | 608 | 127 | 164 | 71 |
| 1980 | 25 | 2636 | 2797 | 729 | 243 | 49 | 55 |
| 1981 | 38 | 1457 | 3635 | 1448 | 244 | 99 | 47 |
| 1982 | 46 | 538 | 2284 | 1455 | 557 | 102 | 79 |
| 1983 | 47 | 1011 | 932 | 751 | 499 | 154 | 46 |
| 1984 | 37 | 1733 | 1195 | 439 | 240 | 161 | 75 |
| 1985 | 34 | 1360 | 2105 | 703 | 158 | 84 | 77 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 49 | 1180 | 2248 | 699 | 203 | 64 | 65 |
| 1987 | 47 | 4522 | 1793 | 841 | 252 | 75 | 43 |
| 1988 | 43 | 2971 | 4734 | 702 | 263 | 71 | 38 |
| 1989 | 41 | 754 | 2163 | 1886 | 231 | 86 | 37 |
| 1990 | 38 | 869 | 1075 | 545 | 372 | 70 | 30 |
| 1991 | 47 | 2169 | 1408 | 442 | 127 | 98 | 22 |
| 1992 | 37 | 1529 | 1243 | 664 | 132 | 42 | 49 |
| 1993 | 39 | 388 | 2907 | 403 | 119 | 16 | 13 |
| 1994 | 40 | 916 | 569 | 848 | 68 | 20 | 10 |
| 1995 | 43 | 678 | 1283 | 180 | 163 | 7 | 6 |
| 1996 | 88 | 447 | 1113 | 700 | 38 | 39 | 6 |
| 1997 | 5 | 651 | 1149.5 | 501 | 213 | 17 | 16 |
| 1998 | 0 | 231 | 1928 | 335 | 80 | 28 | 8 |
| 1999 | 141 | 236 | 843 | 871 | 66 | 21 | 7 |
| 2000 | 62 | 1107 | 176 | 107 | 50 | 4 | 1 |
| 2001 | 7 | 403 | 841 | 53 | 13 | 9 | 2 |
| 2002 | 0 | 238 | 564 | 405 | 7 | 2 | 3 |
| 2003* | 50 | 121 | 472 | 109 | 36 | 1 | 0 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004* | 50 | 161 | 134 | 174 | 22 | 6 | 3 |
| 2005* | 50 | 118 | 256 | 78 | 34 | 5 | 1 |
| 2006 | 50 | 89 | 174 | 128 | 17 | 8 | 3 |
| 2007 | 16 | 216 | 210 | 56 | 11 | 1 | 0 |
| 2008 | 6 | 77 | 169 | 87 | 9 | 3 | 0 |
| 2009 | 329 | 60 | 57 | 66 | 17 | 3 | 0 |
| 2010 | 49 | 220 | 188 | 16 | 7.5 | 2 | 1 |
| 2011 | 10 | 54 | 106 | 36 | 2 | 1 | 1 |
| 2012 | 8 | 84 | 135 | 145 | 10 | 0 | 0 |
| 2013 | 36 | 37 | 59 | 30 | 9 | 2 | 0 |
| 2014 | 1 | 41 | 86 | 26 | 5 | 1 | 0 |
| 2015 | 0 | 37 | 80 | 26 | 4 | 1 | 0 |
| 2016 | 0 | 11 | 25 | 30 | 2 | 1 | 0 |
| 2017 | 0 | 12 | 28 | 16 | 3 | 0 | 0 |
| 2018 | 256 | 95 | 27 | 36 | 2 | 2 | 1 |
| 2019 | 0 | 60 | 68 | 12 | 9 | 1 | 2 |
| 2020* | 0 | 108 | 50 | 20 | 4 | 2 | 1 |
| 2021 | 0 | 11.8 | 22.1 | 13.1 | 4.7 | 0.3 | 0.7 |

*Excluded from assessment as very low sampling.

Table 10. Mean weights-at-age in the landings (used for whole stock and catch). *wean weight at age in landings only available for Q1, hence considerably lower than previous years and only used for forecast.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.86 |
| 1969 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.26 |
| 1970 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.17 |
| 1971 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.12 |
| 1972 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.28 |
| 1973 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.16 |
| 1974 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.34 |
| 1975 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.05 |
| 1976 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.13 |
| 1977 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.63 |
| 1978 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.19 |
| 1979 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.48 |
| 1980 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.87 |
| 1981 | 0.1 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 7.55 |
| 1982 | 0.1 | 1.01 | 1.52 | 3.49 | 5.57 | 7.59 | 9.11 |
| 1983 | 0.1 | 1 | 1.84 | 3.99 | 5.96 | 7.97 | 9.97 |
| 1984 | 0.1 | 0.68 | 1.81 | 3.81 | 5.87 | 7.48 | 10.05 |
| 1985 | 0.1 | 0.78 | 2.02 | 4.24 | 5.83 | 7.5 | 9.04 |
| 1986 | $0.1$ | 0.81 | $1.83$ | $3.86$ | 5.86 | $7.39$ | 8.78 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 0.1 | 0.71 | 2.16 | 3.91 | 6.41 | 7.82 | 10.32 |
| 1988 | 0.1 | 0.61 | 1.56 | 3.76 | 5.67 | 8.02 | 9.88 |
| 1989 | 0.1 | 0.94 | 1.85 | 3.22 | 5.41 | 6.57 | 9.47 |
| 1990 | 0.1 | 0.84 | 1.94 | 3.57 | 5.28 | 7.53 | 9.4 |
| 1991 | 0.1 | 0.86 | 1.64 | 3.54 | 5.42 | 6.39 | 9.11 |
| 1992 | 0.1 | 0.81 | 1.96 | 3.99 | 5.98 | 6.92 | 8.67 |
| 1993 | 0.1 | 0.85 | 1.71 | 3.67 | 5.68 | 7.37 | 10.17 |
| 1994 | 0.1 | 0.8 | 1.92 | 3.61 | 6.08 | 7.68 | 8.57 |
| 1995 | 0.1 | 0.9 | 1.84 | 4.00 | 5.79 | 8.45 | 9.14 |
| 1996 | 0.1 | 0.98 | 1.63 | 3.26 | 5.3 | 7.72 | 9.79 |
| 1997 | 0.1 | 0.85 | 1.94 | 3.62 | 5.29 | 6.12 | 9.4 |
| 1998 | 0.1 | 0.93 | 1.65 | 3.73 | 5.37 | 7.03 | 9.35 |
| 1999 | 0.1 | 0.85 | 1.62 | 3.18 | 5.51 | 7.52 | 10.25 |
| 2000 | 0.1 | 0.85 | 1.99 | 3.57 | 5.14 | 7.15 | 8.39 |
| 2001 | 0.1 | 0.99 | 1.82 | 4.15 | 5.61 | 7.33 | 9.51 |
| 2002 | 0.1 | 0.94 | 1.84 | 3.44 | 5.73 | 7.71 | 10.01 |
| 2003 | 0.1 | 1.21 | 1.66 | 3.29 | 5.43 | 10.2 | 11.09 |
| 2004 | 0.1 | 1.11 | 2.2 | 3.63 | 6.51 | 7.64 | 8.61 |
| 2005 | 0.1 | 0.91 | 1.94 | 3.51 | 5.32 | 7.74 | 8.89 |
| 2006 | 0.1 | 0.83 | 1.84 | 3.67 | 4.71 | $6.39$ | 7.84 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 0.1 | 0.83 | 1.85 | 3.78 | 5.35 | 7.99 | 10.04 |
| 2008 | 0.1 | 0.89 | 1.59 | 3.54 | 6.00 | 7.57 | 9.46 |
| 2009 | 0.1 | 1.1 | 2.01 | 3.46 | 5.31 | 7.1 | 6.82 |
| 2010 | 0.1 | 1.26 | 2.29 | 3.93 | 6.34 | 7.33 | 9.64 |
| 2011 | 0.1 | 0.95 | 1.88 | 3.75 | 5.54 | 6.75 | 9.04 |
| 2012 | 0.1 | 0.93 | 1.88 | 3.37 | 5.34 | 7.60 | 8.56 |
| 2013 | 0.1 | 0.97 | 2.32 | 4.06 | 5.54 | 7.43 | 10.79 |
| 2014 | 0.1 | 0.88 | 2.26 | 4.49 | 7.00 | 8.75 | 9.41 |
| 2015 | 0.1 | 0.83 | 1.79 | 3.69 | 6.49 | 8.55 | 9.95 |
| 2016 | 0.1 | 0.95 | 1.58 | 3.1 | 5.01 | 10.66 | 8.136 |
| 2017 | 0.1 | 0.70 | 1.82 | 3.82 | 5.85 | 7.62 | 9.74 |
| 2018 | 0.1 | 0.43 | 1.69 | 3.64 | 5.56 | 8.58 | 8.70 |
| 2019 | NA | 0.44 | 2.13 | 4.25 | 6.14 | 6.79 | 9.00 |
| 2020 * | 0.1 | 0.22 | 1.29 | 3.67 | 5.23 | 7.85 | 9.54 |
| 2021 | 0.1 | 0.187 | 1.831 | 4.164 | 6.485 | 8.64 | 7.25 |

Table 11. Estimates of numbers discarded (a) and the discarded proportions (b) from 1968-2021. Data are total numbers ('000 fish) discarded at-age, estimated from numbers per sampled trip raised to total fishing effort by each country supplying data (UK, Ireland and Belgium) Please refer to WKIrish3 (ICES, 2017a) documents.
a)

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 17.81 | 74.71 | 0 | 0 | 0 | 0 | 0 |
| 1969 | 20.85 | 87.45 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 22.13 | 92.83 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 22.94 | 96.2 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 26.51 | 111.18 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 27.17 | 113.96 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 16.94 | 71.04 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 26.38 | 110.62 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 26.77 | 112.28 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 31.05 | 130.23 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 39.96 | 167.57 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 44.35 | 185.98 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 24.6 | 103.16 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 37.67 | 157.97 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 46.04 | 193.1 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 46.98 | 197.05 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 37.3 | 156.45 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 33.89 | 142.12 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 49.15 | 206.15 | 0 | 0 | 0 | 0 | 0 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 47.38 | 198.69 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 42.59 | 178.64 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 41.03 | 172.09 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 37.85 | 158.74 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 46.64 | 195.61 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 36.74 | 154.1 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 39.4 | 165.24 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 39.92 | 167.44 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 42.97 | 180.2 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 87.95 | 128.79 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 5.28 | 127.79 | 0.5 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 27.47 | 2 | 0 | 0 | 0 | 0 |
| 1999 | 141.42 | 165.79 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 62.36 | 817.69 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 7.22 | 65.15 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 42.49 | 0 | 0 | 0 | 0 | 0 |
| 2003 * | 50.43 | 75.68 | 32.62 | 15.83 | 1.25 | 0.13 | 0 |
| 2004* | 50.43 | 92.78 | 32.81 | 15.83 | 1.25 | 0.13 | 0 |
| 2005* | 50.43 | 76.34 | 32.36 | 15.83 | 1.25 | 0.13 | 0 |
| 2006 | 50.43 | 75.08 | 32 | 15.83 | 1.25 | 0.13 | 0 |
| 2007 | 16 | 167 | $4.60$ | 0 | 0 | 0 | 0 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 5.50 | 63.40 | 3.40 | 0 | 0 | 0 | 0 |
| 2009 | 329.30 | 39.80 | 4.40 | 0.1 | 0 | 0 | 0 |
| 2010 | 48.70 | 180 | 60.30 | 1.4 | 0.5 | 0.1 | 0 |
| 2011 | 9.70 | 42.70 | 0.90 | 0 | 0 | 0 | 0 |
| 2012 | 7.50 | 79.90 | 100.20 | 112.9 | 5.9 | 0.2 | 0 |
| 2013 | 36.10 | 31 | 26.50 | 11 | 2 | 0.5 | 0 |
| 2014 | 1.09 | 34.66 | 41.93 | 10.3 | 1.53 | 0.1 | 0 |
| 2015 | 0 | 37.30 | 45.80 | 6.8 | 1.3 | 0.3 | 0 |
| 2016 | 0 | 9.84 | 14.15 | 13.45 | 0.91 | 0.74 | 0 |
| 2017 | 0.43 | 9.85 | 7.88 | 8.10 | 0.57 | 0.10 | 0.10 |
| 2018 | 255.50 | 72.19 | 8.89 | 4.88 | 0.12 | 0.22 | 0 |
| 2019 | 0 | 39.2 | 0.4 | 0 | 0 | 0 | 0 |
| 2020* | NA |  |  |  |  |  |  |
| 2021 | 0 | $10.6$ | $6.1$ | $0$ | $0$ | $0$ | 0 |

* very low sampling levels.
b)

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 1 | 0.17 | 0 | 0 | 0 | 0 | 0 |
| 1969 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 1 | 0.05 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 1 | 0.03 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 1 | 0.12 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 1 | 0.05 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 1 | 0.12 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 1 | 0.06 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 1 | 0.14 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 1 | 0.20 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 1 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 1 | 0.04 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 1 | 0.11 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 1 | 0.36 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 1 | 0.19 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 1 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 1 | 0.17 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 1 | 0.04 | $0$ | $0$ | $0$ | $0$ | $0$ |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 1 | 0.06 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 1 | 0.23 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 1 | 0.18 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 1 | 0.09 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 1 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 1 | 0.43 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 1 | 0.18 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 1 | 0.27 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 1 | 0.29 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 1 | 0.20 | 0 | 0 | 0 | 0 | 0 |
| 1998 | NA | 0.12 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 1 | 0.70 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 1 | 0.74 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 1 | 0.16 | 0 | 0 | 0 | 0 | 0 |
| 2002 | NA | 0.18 | 0 | 0 | 0 | 0 | 0 |
| 2003 * | 1 | 0.63 | 0.07 | 0.15 | 0.03 | 0.12 | NA |
| 2004* | 1 | 0.58 | 0.25 | 0.09 | 0.06 | 0.022 | 0 |
| 2005* | 1 | 0.65 | 0.13 | 0.20 | 0.04 | 0.03 | 0 |
| 2006 | 1 | 0.84 | 0.18 | 0.12 | 0.07 | 0.02 | 0 |
| 2007 | 1 | 0.77 | 0.02 | 0 | 0 | 0 | NA |
| 2008 | $1$ | 0.82 | $0.02$ | $0$ | $0$ |  | NA |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 1 | 0.67 | 0.08 | 0 | 0 | 0 | NA |
| 2010 | 1 | 0.82 | 0.32 | 0.06 | 0.07 | 0.05 | 0 |
| 2011 | 1 | 0.80 | 0.01 | 0 | 0 | 0 | 0 |
| 2012 | 1 | 0.95 | 0.74 | 0.78 | 0.60 | 1 | NA |
| 2013 | 1 | 0.84 | 0.45 | 0.37 | 0.22 | 0.34 | NA |
| 2014 | 1 | 0.85 | 0.49 | 0.39 | 0.28 | 0.09 | NA |
| 2015 | NA | 1 | 0.57 | 0.26 | 0.30 | 0.23 | NA |
| 2016 | NA | 0.91 | 0.58 | 0.45 | 0.40 | 0.62 | 0 |
| 2017 | 1 | 0.80 | 0.28 | 0.51 | 0.20 | 0.21 | 0.49 |
| 2018 | 1 | 0.76 | 0.33 | 0.13 | 0.05 | 0.10 | 0 |
| 2019 | NA | 0.65 | <0.01 | 0 | 0 | 0 | 0 |
| 2020* |  |  |  |  |  |  |  |
| 2021 | 1 | 0.89 | $0.28$ | $0$ | $0$ | $0$ | 0 |

NA= not available.
${ }^{*}$ Data for are unavailable due to restricted discard sampling.

| Year | 1 | 2 | 3+ |
| :---: | :---: | :---: | :---: |
| 1996 | 0 | 0.27 | 1 |
| 1997 | 0 | 0.275415 | 1 |
| 1998 | 0 | 0.339514 | 1 |
| 1999 | 0 | 0.402555 | 1 |
| 2000 | 0 | 0.464725 | 1 |
| 2001 | 0 | 0.526111 | 1 |
| 2002 | 0 | 0.585231 | 1 |
| 2003 | 0 | 0.623356 | 1 |
| 2004 | 0 | 0.65373 | 1 |
| 2005 | 0 | 0.676757 | 1 |
| 2006 | 0 | 0.691103 | 1 |
| 2007 | 0 | 0.697111 | 1 |
| 2008 | 0 | 0.700228 | 1 |
| 2009 | 0 | 0.704985 | 1 |
| 2010 | 0 | 0.707035 | 1 |
| 2011 | 0 | 0.704413 | 1 |
| 2012 | 0 | 0.700372 | 1 |
| 2013 | 0 | 0.702394 | 1 |


| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3 +}$ |
| :--- | :--- | :--- | :--- | :--- |
| 2014 | 0 | 0.708485 | 1 |
| 2015 | 0 | 0.716712 | 1 |
| 2016 | 0 | 0.726138 | 1 |
| 2017 | 0 | 0.735987 | 1 |
| 2018 | 0 | 0.745951 | 1 |
| 2019 | 0 | 0.756372 | 1 |
| 2020 | 0 | 0.74887 | 1 |
| 2021 | 0 | 0.75601 | 1 |

## Table 13. Survey catch numbers-at-age and c.v. for all three surveys.

Survey catch numbers-at-age and c.v.

| year | c.v. | 1 | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 0.68 | 700.73 | 386.15 | 20.03 | 10.78 |
| 1996 | 0.42 | 1106.13 | 329.28 | 111.67 | 1.39 |
| 1997 | 0.64 | 537.30 | 415.84 | 66.72 | 21.39 |
| 1998 | 0.84 | 169.36 | 769.23 | 56.87 | 11.98 |
| 1999 | 0.86 | 49.50 | 253.08 | 241.87 | 15.29 |
| 2000 | 0.65 | 629.60 | 101.053 | 34.58 | 33.01 |
| 2001 | 0.89 | 406.68 | 561.44 | 18.44 | 5.78 |
| 2002 | 0.64 | 662.16 | 253.31 | 333.54 | 0 |
| 2003 | 0.54 | 73.87 | 1079.20 | 104.05 | 32.70 |
| 2004 | 0.75 | 216.96 | 171.96 | 88.62 | 5.38 |
| 2005 | 0.76 | 63.53 | 225.07 | 29.41 | 27.96 |
| 2006 | 0.63 | 169.99 | 130.75 | 58.30 | 2.52 |
| 2007 | 0.95 | 164.35 | 124.39 | 30.60 | 5.15 |
| 2008 | 0.90 | 40.66 | 217.15 | 13.02 | 5.17 |
| 2009 | 0.76 | 144.00 | 59.00 | 33.00 | 9.00 |
| 2010 | 0.82 | 1022.12 | 208.96 | 14.66 | 2.26 |
| 2011 | 0.49 | 353.98 | 414.69 | 46.01 | 2.26 |


| year | c.v. | 1 | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 0.81 | 161.90 | 222.82 | 99.27 | 14.25 |
| 2013 | 0.81 | 276.59 | 213.68 | 60.08 | 1.49 |
| 2014 | 0.63 | 314.41 | 222.80 | 53.29 | 13.66 |
| 2015 | 0.84 | 78.96 | 719.35 | 69.19 | 8.56 |
| 2016 | 1.06 | 349.20 | 175.00 | 148.30 | 10.70 |
| 2017 | 0.77 | 69.8 | 445.20 | 57.80 | 12.60 |
| 2018 | 1.26 | 138.1 | 50.50 | 62.60 | 0 |
| 2019 | 0.88 | 214.9 | 171.6 | 27.8 | 14.7 |
| 2020 | 0.977 | 78.5 | 145.4 | 39.4 | 0 |
| 2021 | 1.19 | 86.1 | 158.9 | 38.2 | 0 |

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| year | c.v. | 0 |
| :---: | :---: | :---: |
| 1995 | 0.54163 | 6.66 |
| 1996 | 0.430336 | 12.519 |
| 1997 | 0.720571 | 2.345 |
| 1998 | 0.914513 | 0.047 |
| 1999 | 0.637233 | 6.734 |
| 2000 | 0.785349 | 6.212 |
| 2001 | 0.830289 | 4.863 |
| 2002 | 0.895678 | 0.123 |
| 2003 | 0.707142 | 6.746 |
| 2004 | 0.939137 | 3.663 |
| 2005 | 0.805428 | 8.144 |
| 2006 | 0.871324 | 1.16 |
| 2007 | 1.277817 | 0.067 |
| 2008 | 1.422627 | 0.185 |
| 2009 | 0.938364 | 5.356 |
| 2010 | 1.332794 | 2.779 |
| 2011 | 0.919446 | 0.084 |
| 2012 | 1.256171 | 1.924 |
| 2013 | 0.933411 | 11.208 |
| 2014 | 0.792604 | 0.121 |
| 2015 | 0.872952 | 2.244 |
| 2016 | 1.063181 | 0.149 |
| 2017 | 0.815541 | 4.291 |
| 2018 | 1.419523 | 0.685 |
| 2019 | 1.266571 | 0.072 |
| 2020 | 1.386682 | 0.072 |
| 2021 | 1.610235 | 0.335 |

UK FSP survey

| year | 2 | 3 | 4 | 5 | 6+ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0.43 | 1.41 | 0.99 | 0.08 | 0.03 |
| 2006 | 0.54 | 2.81 | 0.43 | 0.10 | 0.01 |
| 2007 | 0.61 | 1.32 | 0.59 | 0.06 | 0.06 |
| 2008 | 0.22 | 0.82 | 0.15 | 0.08 | 0.02 |
| 2009 | 0.17 | 1.15 | 0.38 | 0.10 | 0.02 |
| 2010 | 0.74 | 0.45 | 0.47 | 0.13 | 0.02 |
| 2011 | 0.41 | 1.68 | 0.14 | 0.10 | 0.04 |
| 2012 | 0.36 | 2.30 | 0.80 | 0.07 | 0.02 |
| 2013 | 0.84 | 1.88 | 1.35 | 0.37 | 0.06 |
| 2014 |  |  |  |  |  |
| 2015 | 0.60 | 2.04 | 1.17 | 0.26 | 0.05 |
| 2016 | 1.00 | 6.39 | 1.43 | 0.41 | 0.03 |
| 2017 | 3.06 | 2.85 | 3.84 | 1.01 | 0.23 |
| 2018 | 0.43 | 3.73 | 0.61 | 0.63 | 0.15 |
| 2019 | 1.30 | 0.75 | 0.83 | 0.12 | 0.19 |
| 2020 | 0.77 | 2.64 | 0.13 | 0.18 | 0.08 |
| 2021 | 0.24 | 0.71 | 0.19 | 0.01 | 0.027 |

Q1 groundfish survey CPUE and SD used in the assessment.

| Year | CPUE | SD |
| :---: | :---: | :---: |
| 1995 | 0.955344 | 0.214285 |
| 1996 | 1.728974 | 0.313405 |
| 1997 | 1.391875 | 0.217769 |
| 1998 | 1.435543 | 0.198929 |
| 1999 | 1.597456 | 0.255936 |
| 2000 | 1.023321 | 0.146161 |
| 2001 | 1.491194 | 0.224681 |
| 2002 | 2.619399 | 0.964573 |
| 2003 | 1.696543 | 0.235312 |
| 2004 | 0.764752 | 0.139312 |
| 2005 | 0.890243 | 0.267329 |
| 2006 | 0.508091 | 0.07914 |
| 2007 | 0.46498 | 0.104631 |
| 2008 | 0.501744 | 0.098636 |
| 2009 | 0.494051 | 0.141257 |
| 2010 | 0.71933 | 0.129658 |
| 2011 | 1.204889 | 0.364965 |
| 2012 | 1.017556 | 0.179033 |
| 2013 | 1.074564 | 0.205801 |
| 2014 | 1.089111 | 0.274391 |
| 2015 | 1.785167 | 0.26655 |
| 2016 | 1.374257 | 0.246976 |
| 2017 | 1.029783 | 0.30429 |
| 2018 | 0.631522 | 0.11959 |
| 2019 | 0.816597 | 0.221725 |
| 2020 | 0.492889 | 0.177333 |
| 2021 | 0.476304 | 0.122131 |

## Table 14. Assessment summary.

| Year | Recruitment | SSB |  |  |  |  | Landings | Discards | Fishing mortality ages 2-4 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Value | High | Low | Value | High | Low |  |  | Value | High | Low |
| 1968 | 184549 | 250405 | 118693 | 46341 | 57748 | 34934 | 8541 | 1285 | 0.108 | 0.156 | 0.059 |
| 1969 | 248327 | 326531 | 170123 | 41512 | 52782 | 30242 | 7991 | 1898 | 0.26 | 0.31 | 0.196 |
| 1970 | 384744 | 489996 | 279492 | 36819 | 47717 | 25921 | 6426 | 708 | 0.28 | 0.33 | 0.24 |
| 1971 | 139637 | 192753 | 86521 | 36214 | 47328 | 25101 | 9246 | 363 | 0.22 | 0.28 | 0.162 |
| 1972 | 358296 | 451004 | 265588 | 41664 | 54267 | 29061 | 9234 | 1546 | 0.26 | 0.33 | 0.194 |
| 1973 | 89533 | 127165 | 51901 | 46927 | 61667 | 32188 | 11819 | 1222 | 0.26 | 0.35 | 0.166 |
| 1974 | 286027 | 363839 | 208215 | 39749 | 52783 | 26715 | 10251 | 1749 | 0.34 | 0.42 | 0.25 |
| 1975 | 95891 | 133523 | 58259 | 40510 | 53818 | 27201 | 9863 | 857 | 0.32 | 0.42 | 0.23 |
| 1976 | 152041 | 202217 | 101865 | 31895 | 43028 | 20762 | 10247 | 381 | 0.33 | 0.44 | 0.23 |
| 1977 | 156311 | 207075 | 105547 | 31257 | 42762 | 19752 | 8054 | 201 | 0.36 | 0.46 | 0.25 |
| 1978 | 292095 | 365791 | 218399 | 24799 | 34579 | 15019 | 5662 | 0 | 0.33 | 0.40 | 0.25 |
| 1979 | 325000 | 401048 | 248952 | 24541 | 33713 | 15368 | 7548 | 0 | 0.25 | 0.34 | 0.162 |
| 1980 | 183575 | 236103 | 131047 | 28154 | 36916 | 19393 | 10599 | 0 | 0.30 | 0.39 | 0.22 |
| 1981 | 87444 | 119000 | 55888 | 36538 | 46259 | 26816 | 13958 | 0 | 0.35 | 0.44 | 0.25 |




| Year | Recruitment |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| age 0 |  |
| Value |  |

*Geometric Mean 2002 to 2019.


Figure 1. Available data.


Figure 2. Landings and discards-at-age. Landings are shaded in grey, discards in white.


Figure 3. Proportion discarded-at-age. Ages 1 and 0 not displayed.


Figure 4. SSB with 95\% confidence interval.


Figure 5. Recruitment with $95 \%$ confidence level. Recruitment in the figure for 2022 is model estimated and not the same as in the forecast.


Figure 6. Age compositions for commercial data and surveys.


Figure 7. Residuals at-age.


Figure 8. Log CPUE fit NIGFS Q1.


Figure 9. Log index fit NIGFS Q4.


Figure 10. Log index fit UKFSP survey.


Figure 11. Results for runs tests for the three indices included and RMSE with fitted LOESS smoother.


Figure 12. Mean age residual fits for total catches, NIGFSQ1 and UKFSP surveys, NIGFSQ4 survey only includes age 0 recruits and is therefore excluded.


Figure 13. Mohns Rho for SSB and $\mathrm{F}_{\text {bar }}$.

